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Algorithm for air quality mapping using satellite images

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1. Introduction

Nowadays, air quality is a major concern in many countries whether in the developed or the developing countries. Environmental pollution is the major concern nowadays because all our daily activities are related to the environmental. Due to the high cost and limited number of air pollutant stations in each area, they cannot provide a good spatial distribution of the air pollutant readings over a city. Satellite observations can give a high spatial distribution of air pollution. The present study is dealing with a new developed algorithm for the determination of the concentration of particulate matter of size less than 10-micron (PM10) over Penang Island, Malaysia.

Air pollution in Asian cities has grown with the progressing industrialization and urbanization. This recent experience in Asia is predated by similar problems in the western countries at early stages of their economic development. Air quality in Chinese cities today is more closely resembles the London smog problem than the Los Angeles smog problem, although that could change as present problems with coal smoke are brought under control (UNEP Assessment Report).

Air pollution causes a number of health problems and it has been linked with illnesses and deaths from heart or lung diseases. Nowadays, air quality is a major problem in many developed countries and they having build up their own network for measuring the air quality levels. Malaysia also has build up our network for monitoring our environment. A network is composed of static measuring stations, which allow continuous measurements of air pollution parameters. Data are collected hourly which include five types of the air pollution constituently such as particulate matter less than 10 micron (PM10), sulphur dioxide (SO₂), nitrogen dioxide (NO₂), carbon monoxide (CO), and ozone (O₃). This network is managed by Alam Sekitar Malaysia Sdn. Bhd. (ASMA), agency contracted by the Department of Environment Malaysia to measure air quality in the country.

Aerosols scatter incoming solar radiation and modify short-wave radiative properties of clouds by acting as cloud condensation nuclei (CCN) (Badarinath, et al.). Particulates matter

(PM), or aerosol, is the general term used for a mixture of solid particles and liquid droplets found in the atmosphere. Monitoring natural (dust and volcanic ash) and anthropogenic aerosols (biomass burning smoke, industrial pollution) has gained renewed attention because they influence cloud properties, alter the radiation budget of the earth-atmosphere system, affect atmospheric circulation patterns and cause changes in surface temperature and precipitation (Wang and Christopher, 2003).

The problem of particulate pollution in the atmosphere has attracted a new interest with the recent scientific evidence of the ill-health effects of small particles. Aerosol optical thickness in the visible (or atmospheric turbidity), which is defined as the linear integral of the extinction coefficient due to small airborne particles, can be considered as an overall air pollution indicator in urban areas. Several studies have shown the possible relationships between satellite data and air pollution. The monitoring of aerosol concentrations becomes a high environmental priority particularly in urban areas. Airborne particulate matter or aerosols, whether anthropogenic or of natural origin constitute a major environmental issue. At the regional level, aerosols are contributors to visibility degradation (haze) and to acid deposition; at the global level they play a role in climate change. At local level, epidemiological studies indicate that small sized aerosols are causal factors in mortality in urban areas. Therefore monitoring aerosol concentrations becomes a high priority at a variety of geographical scales.

Remote sensing technique was wide used for environment pollutant application such as water quality [Dekker, et al., (2002), Tassan, (1993) and Doxaran, et al., (2002)] and air pollutant (Ung, et al., 2001b). Several studies have shown the possible relationships between satellite data and air pollution [Weber, et al., (2001) and Ung, et al., (2001a)]. Other researchers used satellite data for such environment atmospheric studies such as NOAA-14 AVHRR (Ahmad and Hashim, 1997) and Landsat TM (Ung, et al., 2001b). In fact, air quality can be measure using ground instrument such as air sample. But these instruments are quite expensive and the coverage is limited by the number of the air pollutant station in each area. So, they cannot provide a good spatial distribution of air pollutant readings over a city. Compared to ground measurements, satellite imagery, due to their large spatial coverage and reliable repeated measurements, provide another important tool to monitor aerosols and their transport patterns (Wang and Christopher, 2003). The atmosphere affects satellite images of the Earth's surface in the solar spectrum. So, the signal observe by the satellite sensor was the sum of the effects from the ground and atmosphere. Tropospheric aerosols act to significantly alter the Earth's radiation budget, but quantification of the change in radiation is difficult because atmospheric aerosol distributions vary greatly in type, size, space and time (Penner, et al. 2002). Surface reflectance is a key to the retrieval of atmospheric components from remotely sensed data. Optical atmospheric effects may influence the signal measured by a remote sensor in two ways: radiometrically and geometrically. This means that they can modify the signal's intensity through scattering or absorption processes and its direction by refraction (Sifakis and Deschamps, 1992).

The objective of this study was to evaluate the developed algorithm for PM₁₀ mapping by using Landsat visible bands. The corresponding PM₁₀ data were measured simultaneously with the acquisition satellite scene for algorithm regression analysis. The algorithm was developed based on the atmospheric optical model. An algorithm was developed for PM₁₀ determination. The independent variables are the visible wavelengths reflectance signals.

2. Study Area

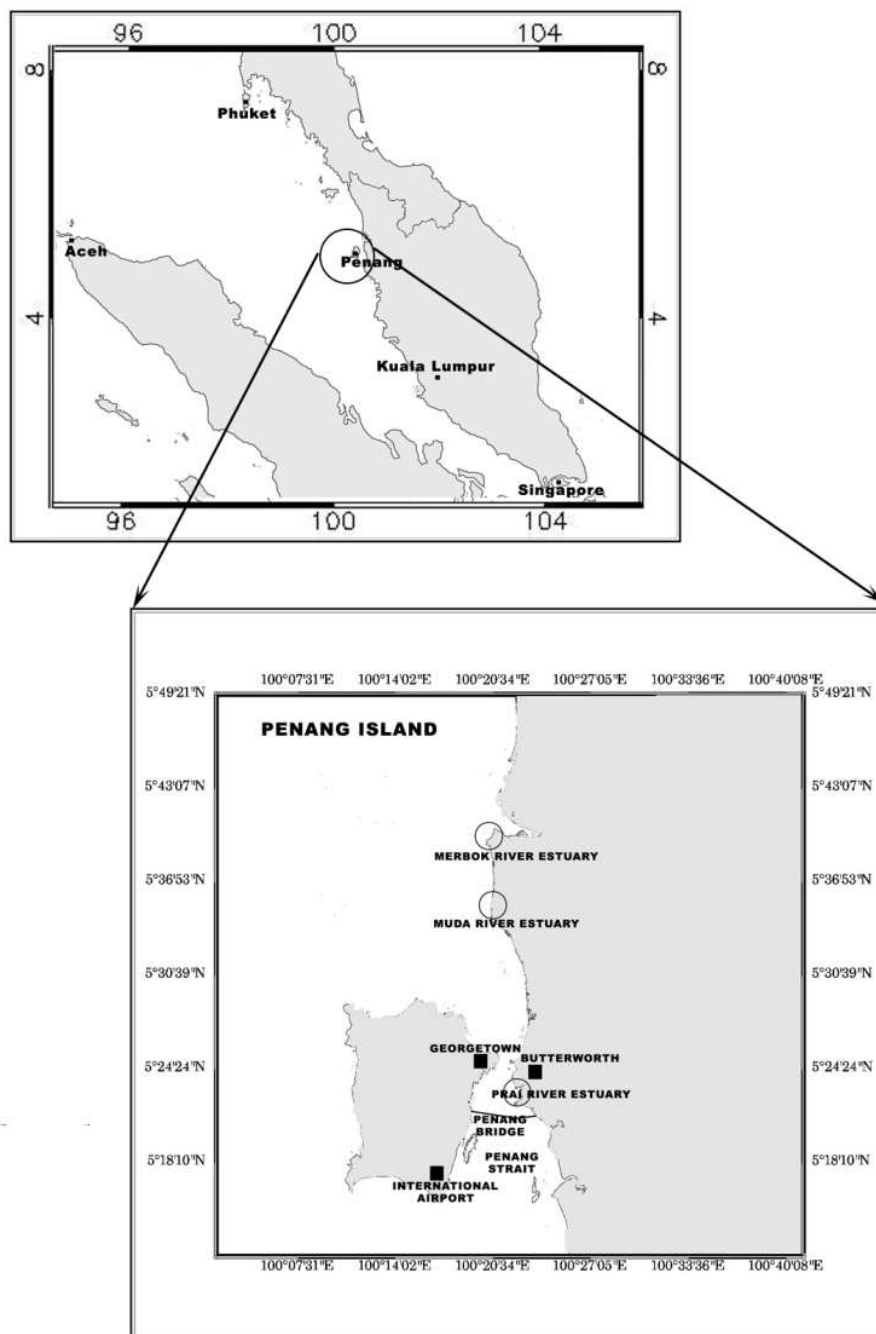


Fig. 1. Study area.

The study area is the Penang Island, Malaysia, located within latitudes $5^{\circ} 9' N$ to $5^{\circ} 33' N$ and longitudes $100^{\circ} 09' E$ to $100^{\circ} 30' E$ (Figure 1). The map of the study area is shown in Figure 1. Penang Island is located in equatorial region and enjoys a warm equatorial weather the whole year. Therefore, it is impossible to get the 100 % cloud free satellite image over Penang Island. But, the satellite image chosen is less than 10 % of cloud coverage over the study area. Penang Island located on the northwest coast of Peninsular Malaysia.

Penang is one of the 13 states of the Malaysia and the second smallest state in Malaysia after Perlis. The state is geographically divided into two different entities - Penang Island (or "Pulau Pinang" in Malay Language) and a portion of mainland called "Seberang Perai" in Malay Language. Penang Island is an island of 293 square kilometres located in the Straits of Malacca and "Seberang Perai" is a narrow hinterland of 753 square kilometres (Penang-Wikipedia, 2009). The island and the mainland are linked by the 13.5 km long Penang Bridge and ferry.

Penang Island is predominantly hilly terrain, the highest point being Western Hill (part of Penang Hill) at 830 metres above sea level. The terrain consists of coastal plains, hills and mountains. The coastal plains are narrow, the most extensive of which is in the northeast which forms a triangular promontory where George Town, the state capital, is situated. The topography of "Seberang Perai" is mostly flat. Butterworth, the main town in "Seberang Perai", lies along the "Perai" River estuary and faces George Town at a distance of 3 km (2 miles) across the channel to the east (Penang-Wikipedia, 2009).

The Penang Island climate is tropical, and it is hot and humid throughout the year. with the average mean daily temperature of about 27°C and mean daily maximum and minimum temperature ranging between 31.4°C and 23.5°C respectively. However, the individual extremes are 35.7°C and 23.5°C respectively. The mean daily humidity varies between 60.9% and 96.8%. The average annual rainfall is about 267 cm and can be as high as 624 cm (Fauziah, et al, 2006).

3. Algorithm Model

The atmospheric reflectance due to molecule, R_r , is given by (Liu, et al., 1996)

$$R_r = \frac{\tau_r P_r(\Theta)}{4\mu_s \mu_v} \quad (1)$$

where

τ_r = Aerosol optical thickness (Molecule)

$P_r(\Theta)$ = Rayleigh scattering phase function

μ_v = Cosine of viewing angle

μ_s = Cosine of solar zenith angle

We assume that the atmospheric reflectance due to particle, R_a , is also linear with the τ_a [King, et al., (1999) and Fukushima, et al., (2000)]. This assumption is reasonable because other researchers also found the linear relationship between both aerosol and molecule scattering (Liu, et al., 1996).

$$R_a = \frac{\tau_a P_a(\Theta)}{4\mu_s \mu_v} \quad (2)$$

where

τ_a = Aerosol optical thickness (aerosol)

$P_a(\Theta)$ = Aerosol scattering phase function

Atmospheric reflectance is the sum of the particle reflectance and molecule reflectance, R_{atm} , (Vermote, et al., (1997).

$$R_{atm} = R_a + R_r \quad (3)$$

where

R_{atm} = atmospheric reflectance

R_p = particle reflectance

R_r = molecule reflectance

$$R_{atm} = \left[\frac{\tau_a P_a(\Theta)}{4\mu_s \mu_v} + K_1 + \frac{\tau_r P_r(\Theta)}{4\mu_s \mu_v} \right]$$

$$R_{atm} = \frac{1}{4\mu_s \mu_v} [\tau_a P_a(\Theta) + \tau_r P_r(\Theta)] \quad (4)$$

The optical depth is given by other researcher as in equation (5) (Camagni and Sandroni, 1983). From the equation, we rewrite the optical depth for particle and molecule as

$$\tau = \sigma \rho s \quad (5)$$

where

τ = optical depth

σ = absorption

s = finite path

$$\tau = \tau_a + \tau_r \quad (\text{Camagni and Sandroni, 1983})$$

$$\tau_r = \sigma_r \rho_r s \quad (6a)$$

$$\tau_p = \sigma_p \rho_p s \quad (6b)$$

Equations (6) are substituted into equation (4). The result was extended to a three bands algorithm as equation (8).

Form the equation; we found that PM10 was linearly related to the reflectance for band 1 and band 2. This algorithm was generated based on the linear relationship between τ and reflectance. Other Researcher also found that the PM10 was linearly related to the τ and the correlation coefficient for linear was better than exponential in their study (overall) (Retalis et al., 2003). This means that reflectance was linear with the PM10. In order to simplify the data processing, the air quality concentration was used in our analysis instead of using density, ρ , values.

$$R_{atm} = \frac{1}{4\mu_s \mu_v} [\sigma_a \rho_a s P_a(\Theta) + \sigma_r \rho_r s P_r(\Theta)]$$

$$R_{atm} = \frac{s}{4\mu_s \mu_v} [\sigma_a \rho_a P_a(\Theta) + \sigma_r \rho_r P_r(\Theta)]$$

$$R_{atm}(\lambda_1) = \frac{S}{4\mu_s\mu_v} [\sigma_a(\lambda_1)PP_a(\Theta, \lambda_1) + \sigma_r(\lambda_1)GP_r(\Theta, \lambda_1)]$$

$$R_{atm}(\lambda_2) = \frac{S}{4\mu_s\mu_v} [\sigma_a(\lambda_2)PP_a(\Theta, \lambda_2) + \sigma_r(\lambda_2)GP_r(\Theta, \lambda_2)]$$

$$P = a_0 R_{atm}(\lambda_1) + a_1 R_{atm}(\lambda_2) \quad (7)$$

The equation (8) was for two bands, so we rewrite the PM10 equation in three bands as

$$P = a_0 R_{atm}(\lambda_1) + a_1 R_{atm}(\lambda_2) + a_2 R_{atm}(\lambda_3) \quad (8)$$

Where

P = Particle concentration (PM10)

G = Molecule concentration

R_{atm*i*} = Atmospheric reflectance, i = 0, 1 and 3 are the number of the band

e_j = algorithm coefficient, j = 0, 1, 2, ... (empirically determined).

4. Data Analysis and Results

Remote sensing satellite detectors exhibit linear response to incoming radiance, whether from the Earth's surface radiance or internal calibration sources. This response is quantized into 8-bit values that represent brightness values commonly called Digital Numbers (DN). To convert the calibrated digital numbers to at-aperture radiance, rescaling gains and biases are created from the known dynamic range limits of the instrument.

$$\text{Radiance, } L(\lambda) = \text{Bias}(\lambda) + [\text{Gain}(\lambda) \times \text{DN}(\lambda)] \quad (9)$$

where

λ = band number.

L is the radiance expressed in $\text{Wm}^{-2} \text{sr}^{-1} \mu\text{m}^{-1}$.

The spectral radiance, as calculated above, can be converted to at sensor reflectance values.

$$\rho^* = \frac{\pi L(\lambda) d^2}{E_0(\lambda) \cos \theta} \quad (10)$$

Where

ρ^* = Sensor Reflectance values

L(λ) = Apparent At-Sensor Radiance ($\text{Wm}^{-2} \text{sr}^{-1} \mu\text{m}^{-1}$)

d = Earth-Sun distance in astronomical units

= $\{1.0 - 0.016729 \cos [0.9856(D-4)]\}$ where (D = day of the year)

E⁰(λ) = mean solar exoatmospheric irradiance ($\text{Wm}^{-2} \text{sr}^{-1} \mu\text{m}^{-1}$)

θ = solar Zenith angle (degrees)

The retrieval of surface reflectance is important to obtain the atmospheric reflectance in remotely sensed data and later used for algorithm calibration. In this study, we retrieve the surface reflectance using the relationship between the two visible bands (blue and red) and the mid infrared data at 2.1 μm . We use the assumption that the mid infrared band data is not significantly affected by atmospheric haze. An algorithm was developed based on the aerosol properties to correlate the atmospheric reflectance and PM10.

The surface reflectance can be obtained from mid-infrared band because the surface reflectances at various bands across the solar spectrum are correlated to each other to some extent. The surface reflectances of dark targets in the blue and red bands were estimated using the measurements in the mid-infrared band (Quaidrari and Vermote, 1999). Over a simple black target, the observed atmospheric reflectance is the sum of reflectance of aerosols and Rayleigh contributions (Equation 9). This simplification, however, is not valid at short wavelengths (less than 0.45 μm) or large sun and view zenith angles (Vermote and Roger, 1996). In this study, a simple form of the equation was used in this study (Equation 11). This equation also used by other research in their study (Popp, 2004).

$$\begin{aligned} R_s - T R_r &= R_{\text{atm}} & (11) \\ R_s - R_r &= R_{\text{atm}} & (12) \end{aligned}$$

where:

R_s = reflectance recorded by satellite sensor

R_r = reflectance from surface references

R_{atm} = reflectance from atmospheric components (aerosols and molecules)

T = transmittance

The surface reflectance in mid-infrared band is related to those in the visible bands. The surface reflectances were measured using a handheld spectroradiometer in the wavelength range of visible wavelengths (red and blue bands). The surface reflectances of dark targets in the visible bands are as follows:

$$\begin{aligned} \rho(\text{TM1}) &= \rho(\text{TM7})/3.846 \\ \rho(\text{TM3}) &= \rho(\text{TM7})/1.923 \end{aligned} \quad (13)$$

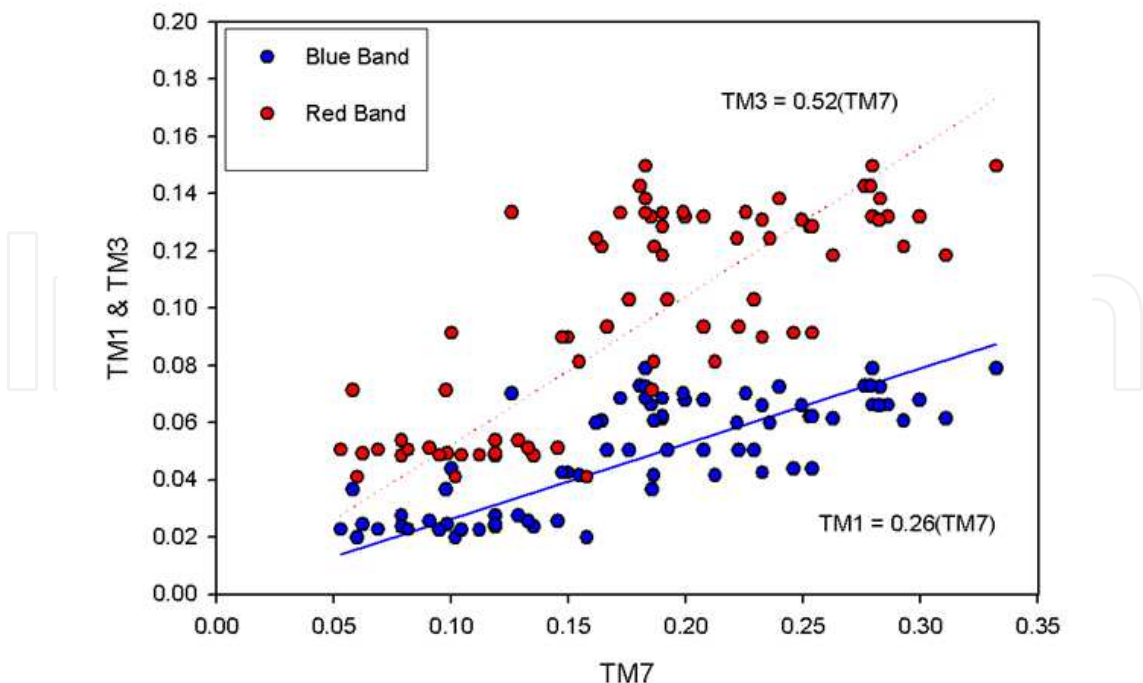


Fig. 2. The relationship between band 1 (TM1) and band 3 (TM3) with band 7 (TM7) of Landsat.

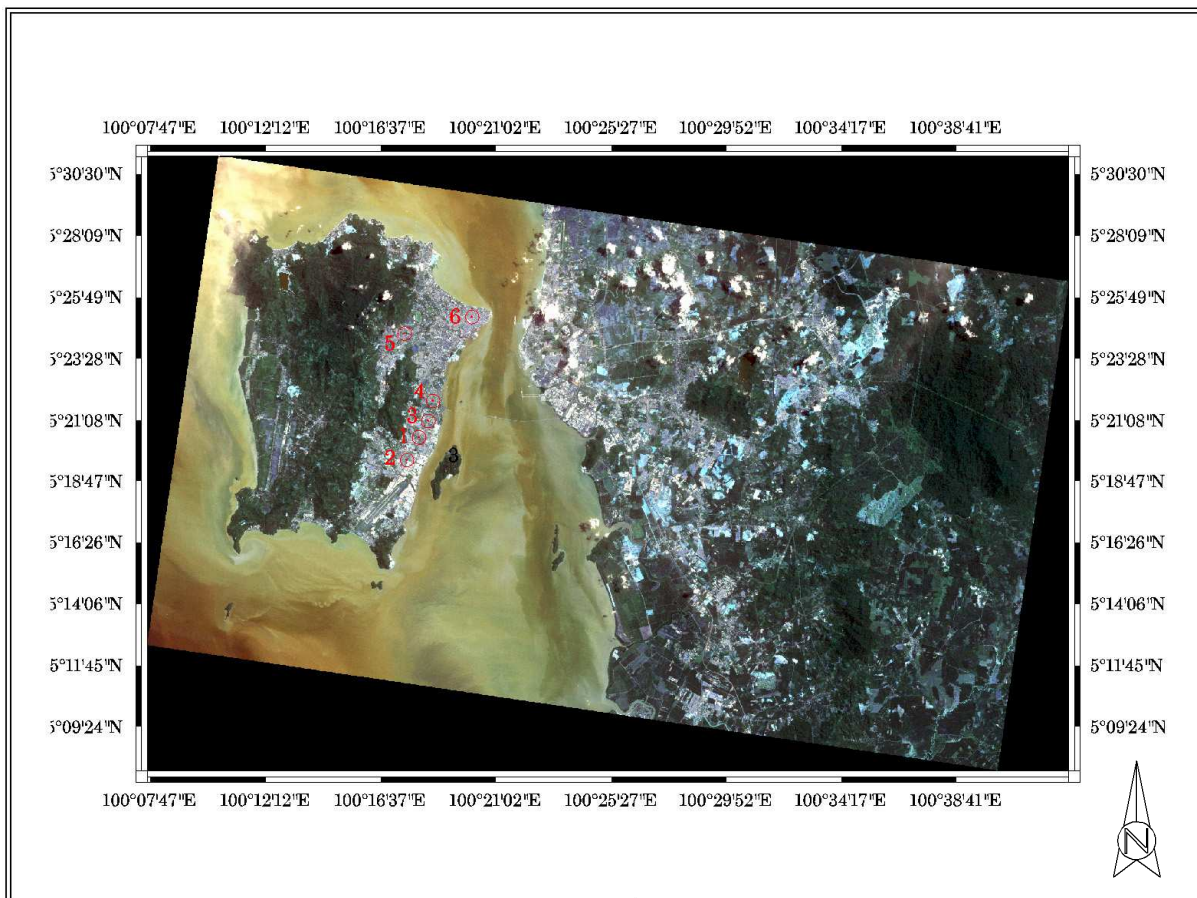


Fig. 3. Raw Landsat TM satellite image of 30/7/2000

Landsat satellite data set was selected corresponding to the ground truth measurements of the pollution levels. The PCI Geomatica version 10.1 image processing software was used in all the analyses. The Landsat satellite images were acquired on 30th July 2000 (Figure 3), 15th February 2001 (Figure 4), 17th January 2002 (Figure 5), 6th March 2002 (Figure 6), 5 February 2003 (Figure 7), 19th March 2004 (Figure 8) and 2nd February 2005 (Figure 9).

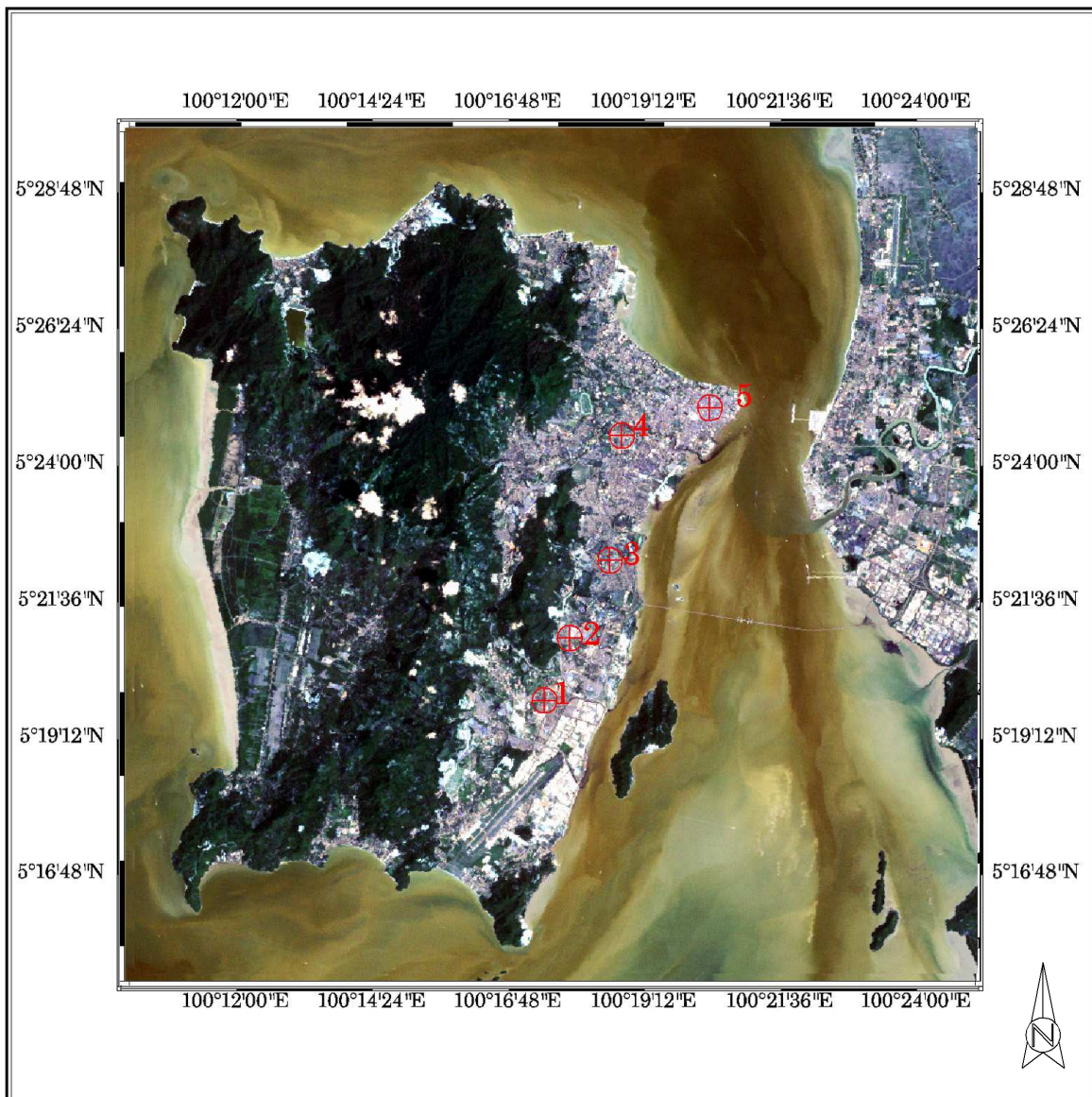


Fig. 4. Raw Landsat TM satellite image of 15/2/2001

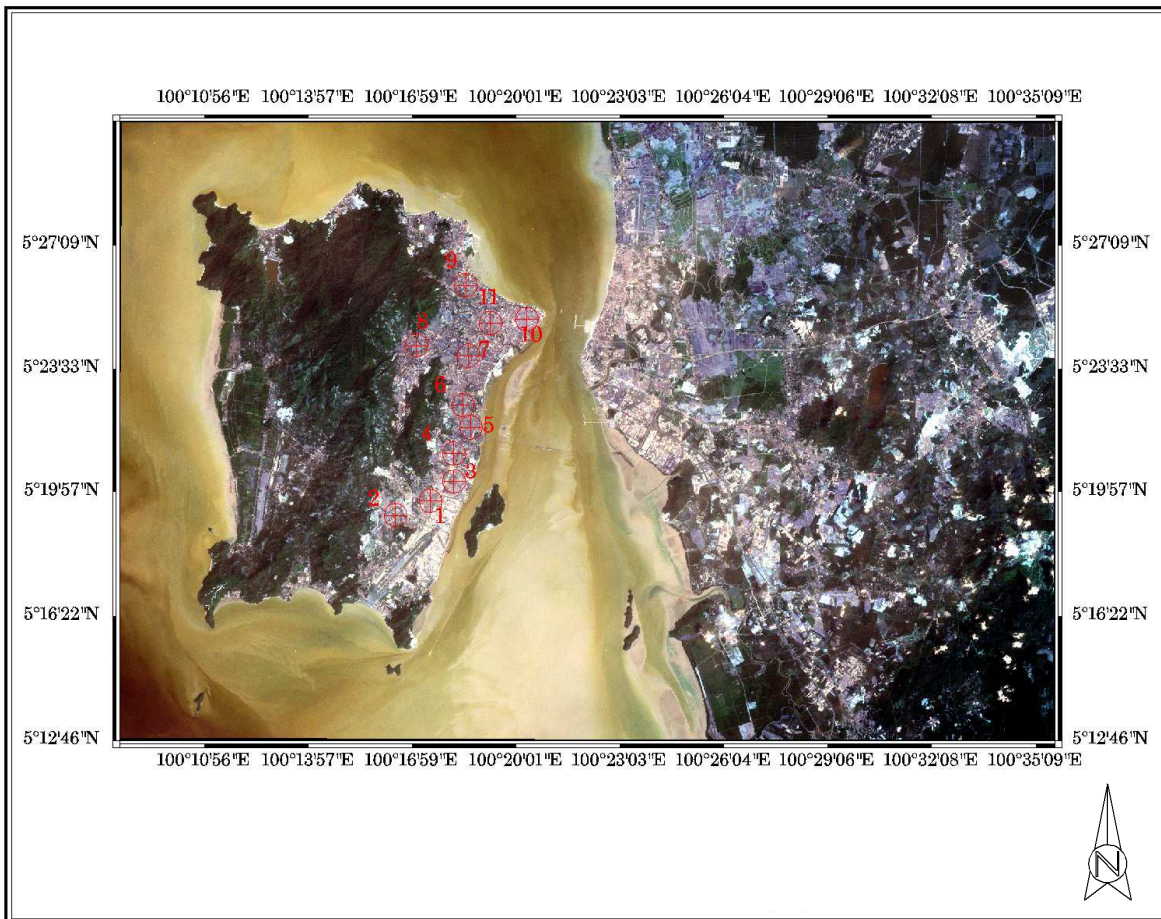


Fig. 5. Raw Landsat TM satellite image of 17/1/2002

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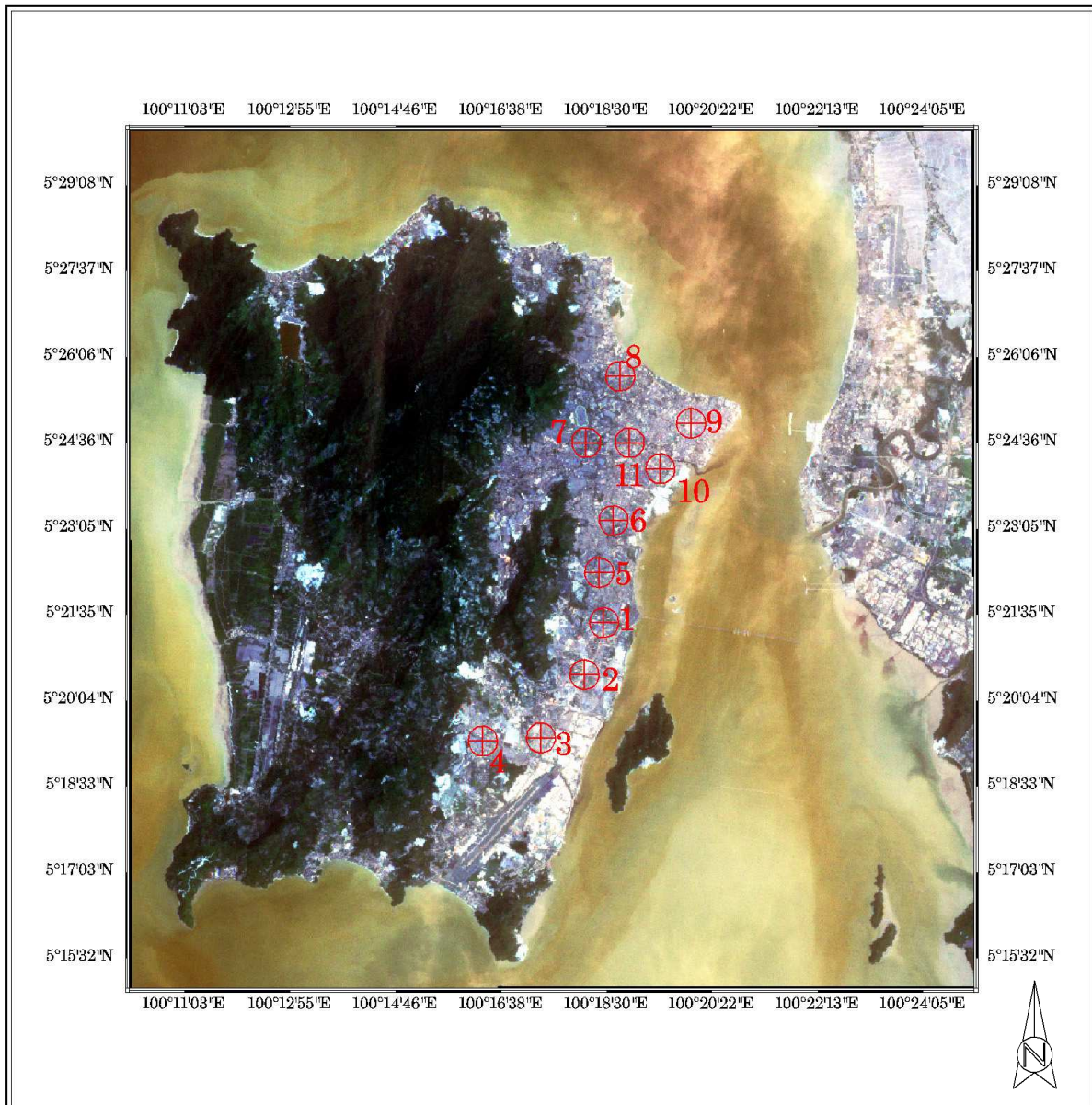


Fig. 6. Raw Landsat TM satellite image of 6/3/2002

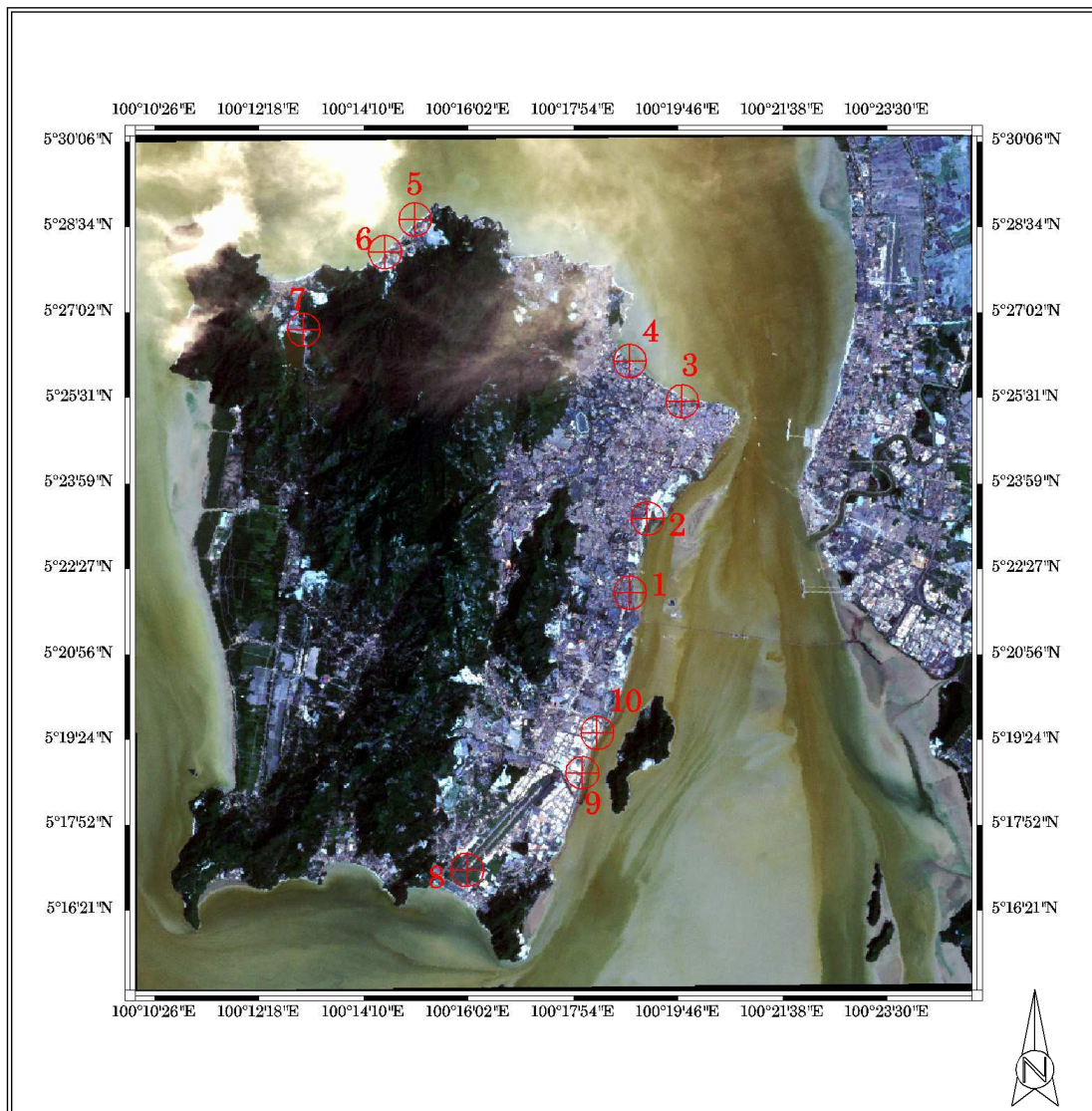


Fig. 7. Raw Landsat TM satellite image of 5/2/2003

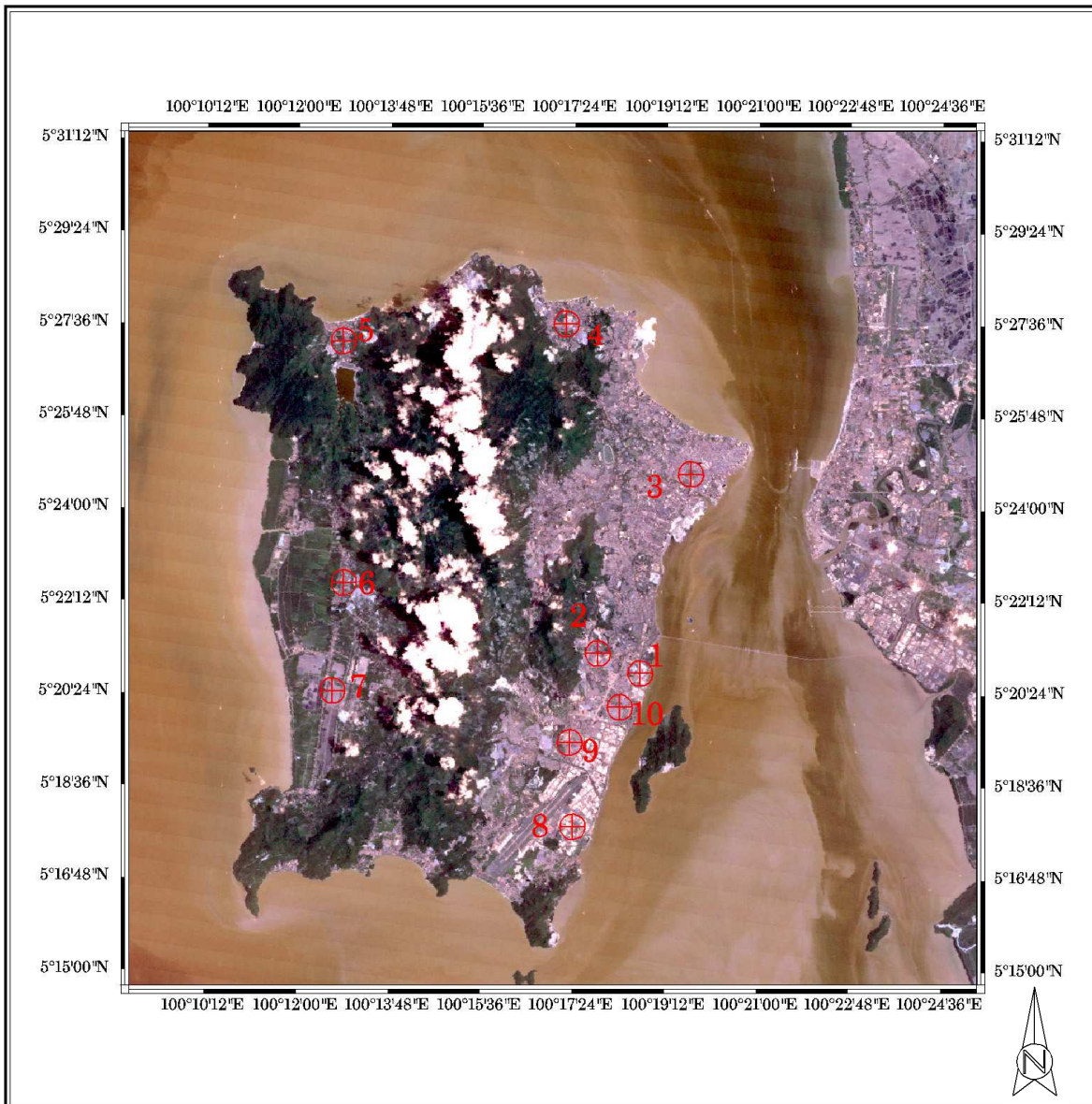


Fig. 8. Raw Landsat TM satellite image of 19/3/2004

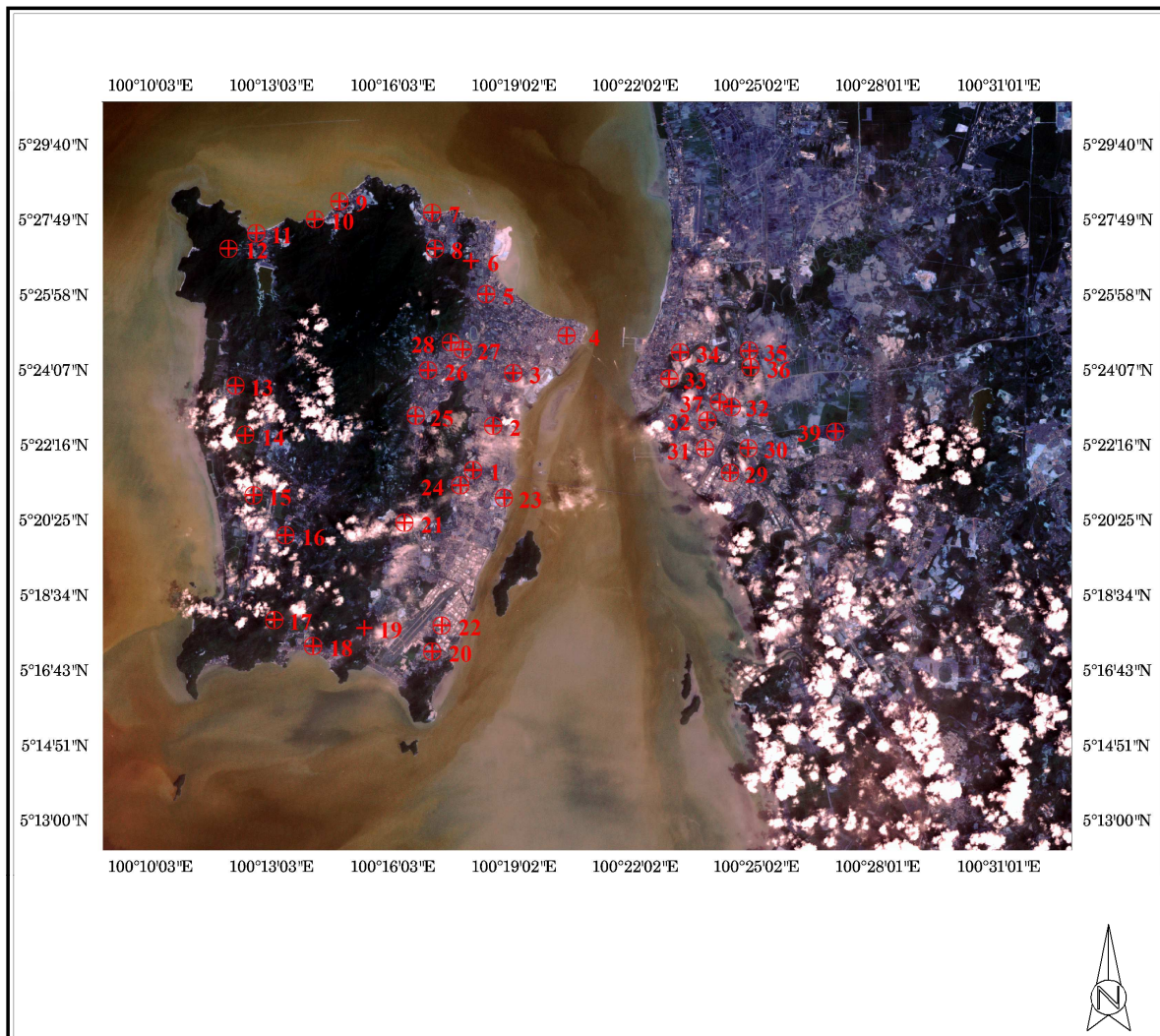


Fig. 9. Raw Landsat TM satellite image of 2/2/2005

Raw digital satellite images usually contain geometric distortion and cannot be used directly as a map. Some sources of distortion are variation in the altitude, attitude and velocity of the sensor. Other sources are panoramic distortion, earth curvature, atmospheric refraction and relief displacement. So, to correct the images, we have to do geometric correction. Image rectification was performed by using a second order polynomial transformation equation. The images were geometrically corrected by using a nearest neighbour resampling technique. Sample locations were then identified on these geocoded images. Regression technique was employed to calibrate the algorithm using the satellite multispectral signals.

PM₁₀ measurements were collected simultaneously with the image acquisition using a DustTrak Aerosol Monitor 8520. The digital numbers of the corresponding in situ data were converted into irradiance and then reflectance. Our approach to retrieve the atmospheric component from satellite observation is by measuring the surface component properties. The reflectance measured from the satellite [reflectance at the top of atmospheric, $\rho(\text{TOA})$] was subtracted by the amount given by the surface reflectance to obtain the atmospheric

reflectance. And then the atmospheric reflectance was related to the PM10 using the regression algorithm analysis. For each visible band, the dark target surface reflectance was estimated from that of the mid-infrared band. The atmospheric reflectance values were extracted from the satellite observation reflectance values subtracted by the amount given by the surface reflectance. The atmospheric reflectance were determined for each band using different window sizes, such as, 1 by 1, 3 by 3, 5 by 5, 7 by 7, 9 by 9 and 11 by 11. In this study, the atmospheric reflectance values extracted using the window size of 3 by 3 was used due to the higher correlation coefficient (R) with the ground-truth data.

The atmospheric reflectance values for the visible bands of TM1 and TM3 were extracted corresponding to the locations of in situ PM10 data. The relationship between the reflectance and the corresponding air quality data was determined using regression analysis. A new algorithm was developed for detecting air pollution from the digital images chosen based on the highest correlation coefficient, R and lowest root mean square error, RMS for PM10. The algorithm was used to correlate atmospheric reflectance and the PM10 values. The proposed algorithm produced high correlation coefficient (R) and low root-mean-square error (RMS) between the measured and estimated PM10 values. Finally, PM10 maps were generated using the proposed algorithm. This study indicates the potential of Landsat for PM10 mapping.

The data points were then regressed to obtain all the coefficients of equation (8). Then the calibrated algorithm was used to estimate the PM10 concentrated values for each image. The proposed model produced the correlation coefficient of 0.83 and root-mean-square error 18 $\mu\text{g}/\text{m}^3$. The PM10 maps were generated using the proposed calibrated algorithm. The generated PM10 map was colour-coded for visual interpretation [Figures 10 - 16]. Generally, the concentrations above industrial and urban areas were higher compared to other areas.

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Algoritma	R				
	S1	S2	S3	S4	S5
$PM10=a_0+a_1B_1+a_2B_1^2$	0.8670	0.8828	0.4893	0.6630	0.8596
$PM10=a_0+a_1B_3+a_2B_3^2$	0.8773	0.9434	0.8415	0.7083	0.8884
$PM10=a_0+a_1\ln B_1+a_2(\ln B_1)^2$	0.9196	0.8944	0.4860	0.6293	0.8698
$PM10=a_0+a_1\ln B_3+a_2(\ln B_3)^2$	0.8897	0.9416	0.8418	0.7108	0.8954
$PM10=a_0+a_1(B_1/B_3)+a_2(B_1/B_3)^2$	0.5655	0.8078	0.2038	0.4039	0.7896
$PM10=a_0+a_1\ln(B_1/B_3)+a_2\ln(B_1/B_3)^2$	0.6494	0.8052	0.1676	0.3431	0.7955
$PM10=a_0+a_1(B_1-B_3)+a_2(B_1-B_3)^2$	0.2663	0.1737	0.6507	0.3281	0.6903
$PM10=a_1B_1+a_2B_3$ (Dicadangkan)	0.9250	0.9520	0.8834	0.8890	0.9042

* B_1 and B_3 are the atmospheric reflectance values for red, green and blue band respectively.

Table 1 Regression results (R) using different forms of algorithms for PM10

Algoritma	RMS ($\mu\text{g}/\text{m}^3$)				
	S1	S2	S3	S4	S5
$PM10=a_0+a_1B_1+a_2B_1^2$	10.6062	5.7532	13.5174	12.3583	8.7407
$PM10=a_0+a_1B_3+a_2B_3^2$	10.2125	4.0631	8.4278	11.6537	7.8532
$PM10=a_0+a_1\ln B_1+a_2(\ln B_1)^2$	8.3605	5.4773	13.5726	12.8299	8.4424
$PM10=a_0+a_1\ln B_3+a_2(\ln B_3)^2$	9.7171	4.1251	8.4115	11.6123	7.6172
$PM10=a_0+a_1(B_1/B_3)+a_2(B_1/B_3)^2$	17.5531	7.2187	16.7673	15.1016	10.4991
$PM10=a_0+a_1\ln(B_1/B_3)+a_2\ln(B_1/B_3)^2$	16.1839	7.2633	16.9753	15.5062	10.3644
$PM10=a_0+a_1(B_1-B_3)+a_2(B_1-B_3)^2$	20.5137	12.0613	11.0887	15.5941	12.3781
$PM10=a_1B_1+a_2B_3$ (Dicadangkan)	9.9045	5.3033	9.2470	8.0795	7.3062

* B_1 and B_3 are the atmospheric reflectance values for red, green and blue band respectively.

Table 2 Regression results (RMS) using different forms of algorithms for PM10

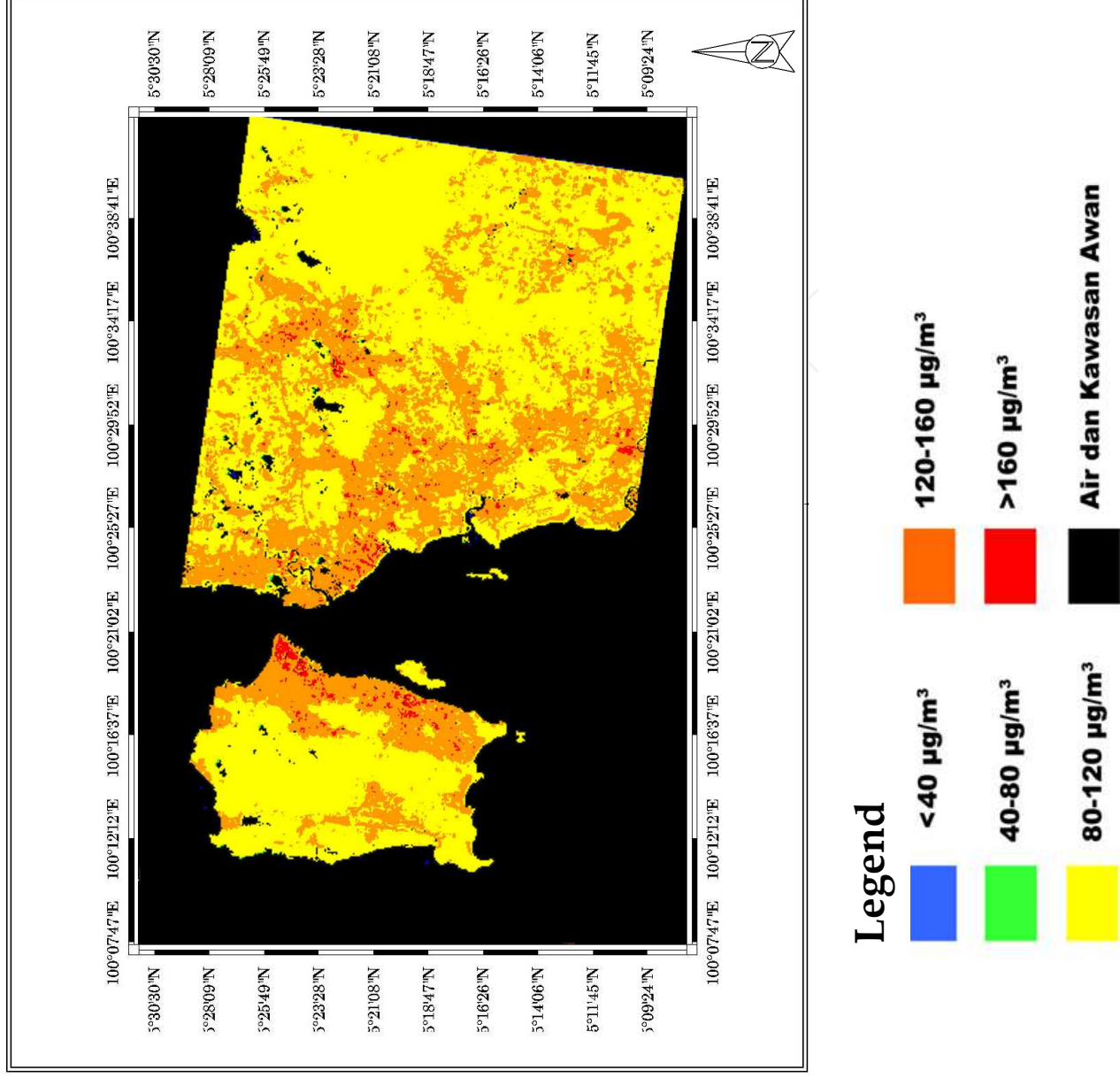


Fig. 10. Map of PM10 around Penang Island, Malaysia-30/7/2000

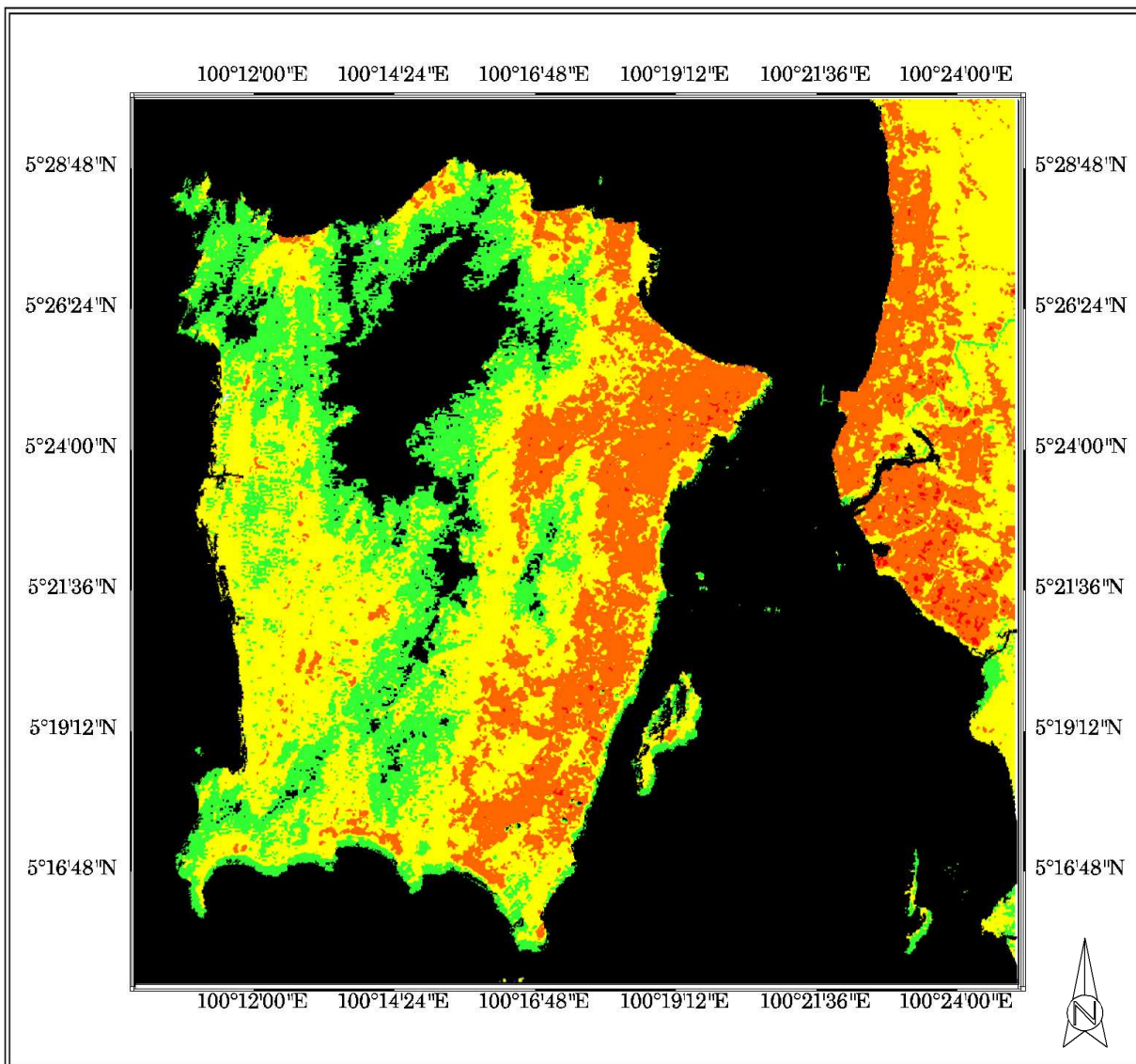
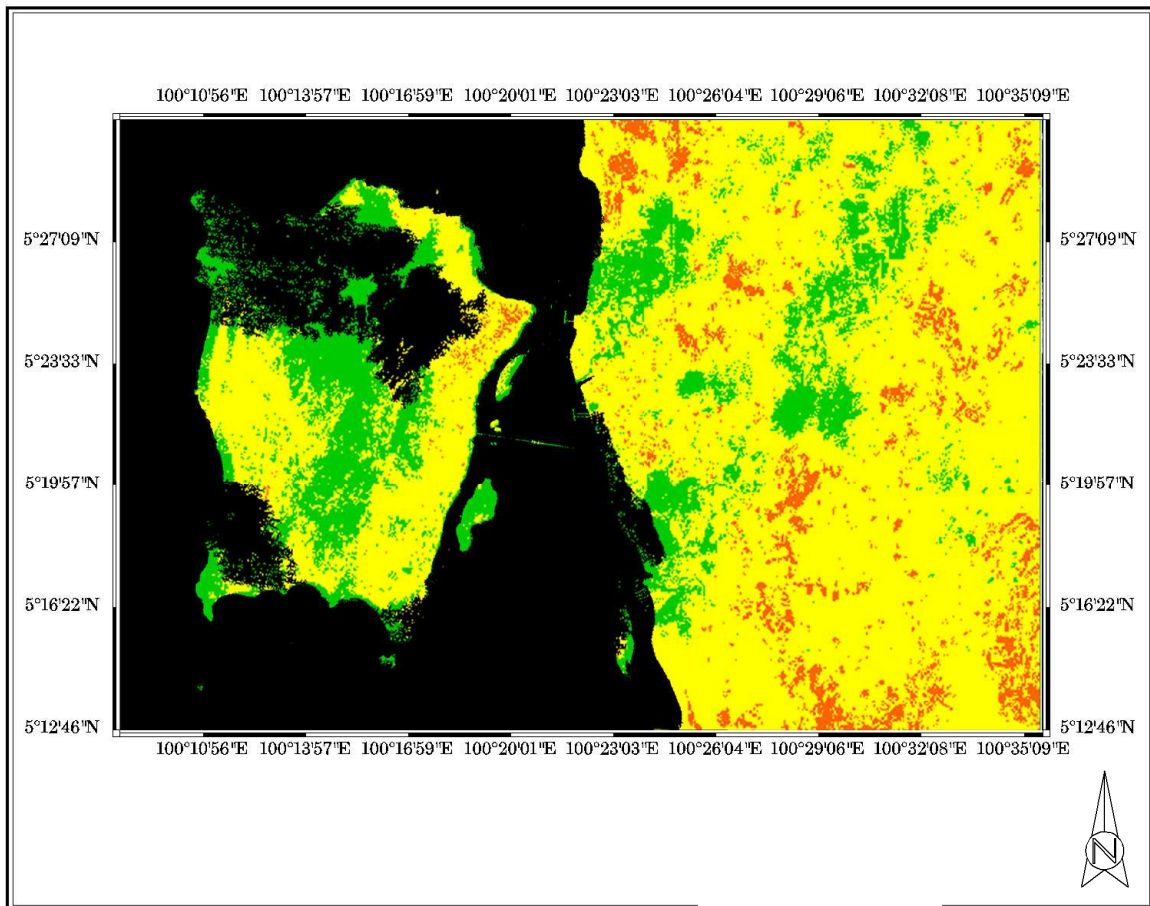


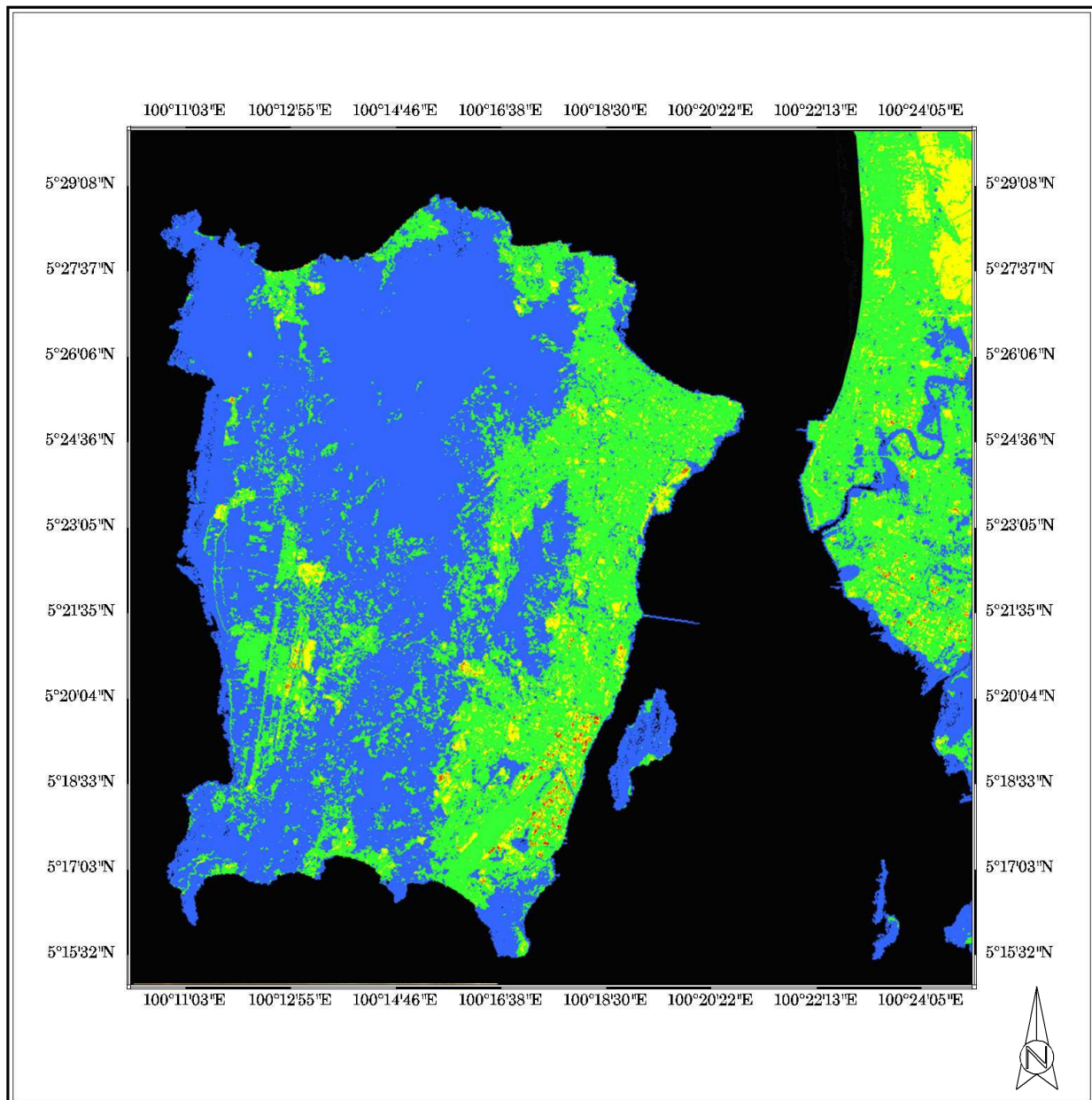
Fig. 11. Map of PM₁₀ around Penang Island, Malaysia-15/2/2001



Legend



Fig. 12. Map of PM10 around Penang Island, Malaysia-17/1/2002



Legend



Fig. 13. Map of PM10 around Penang Island, Malaysia-6/3/2002

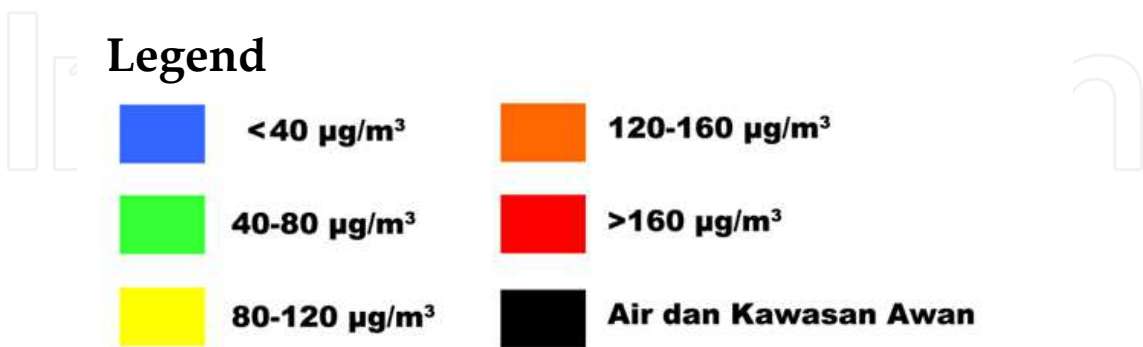
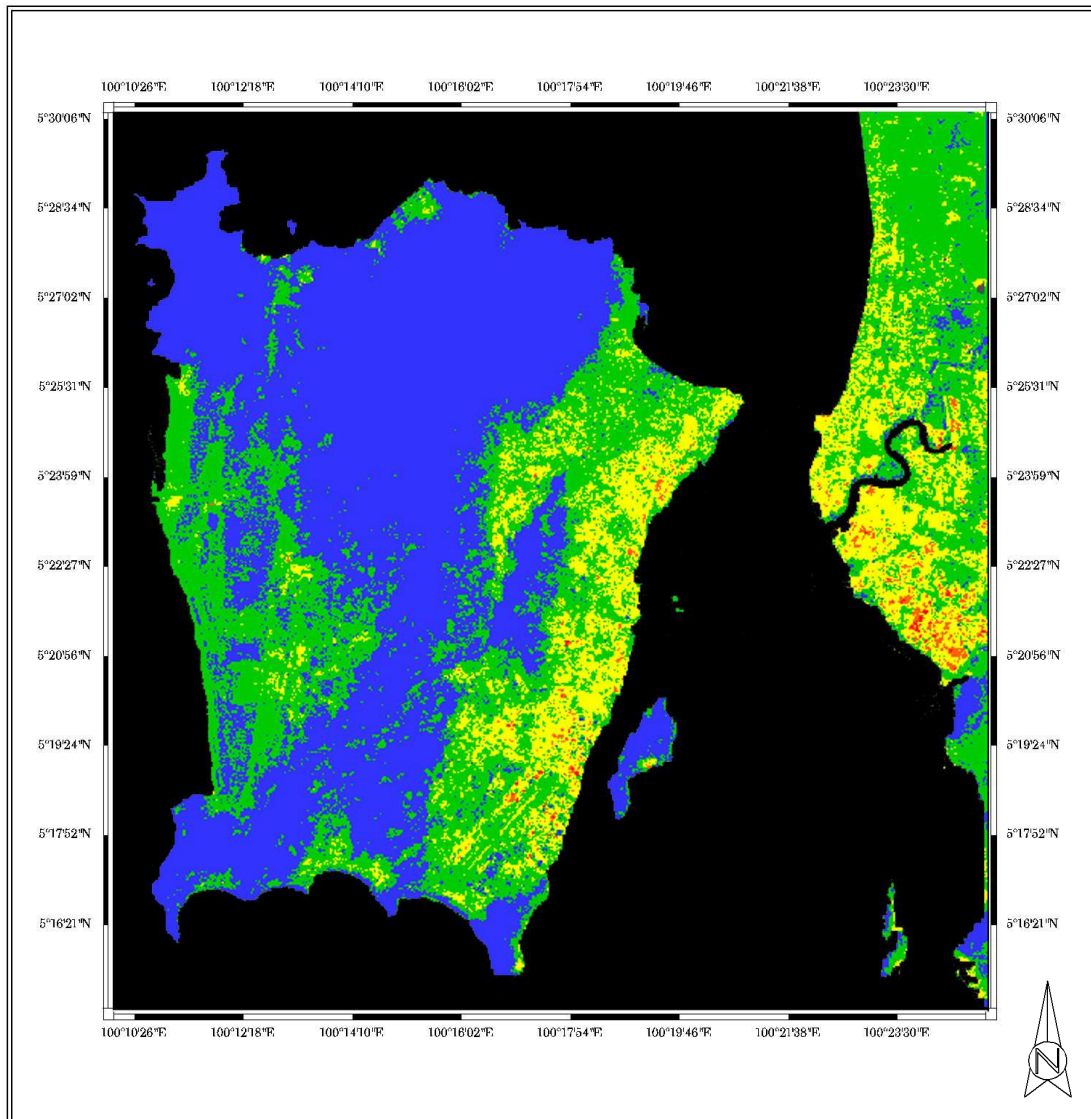
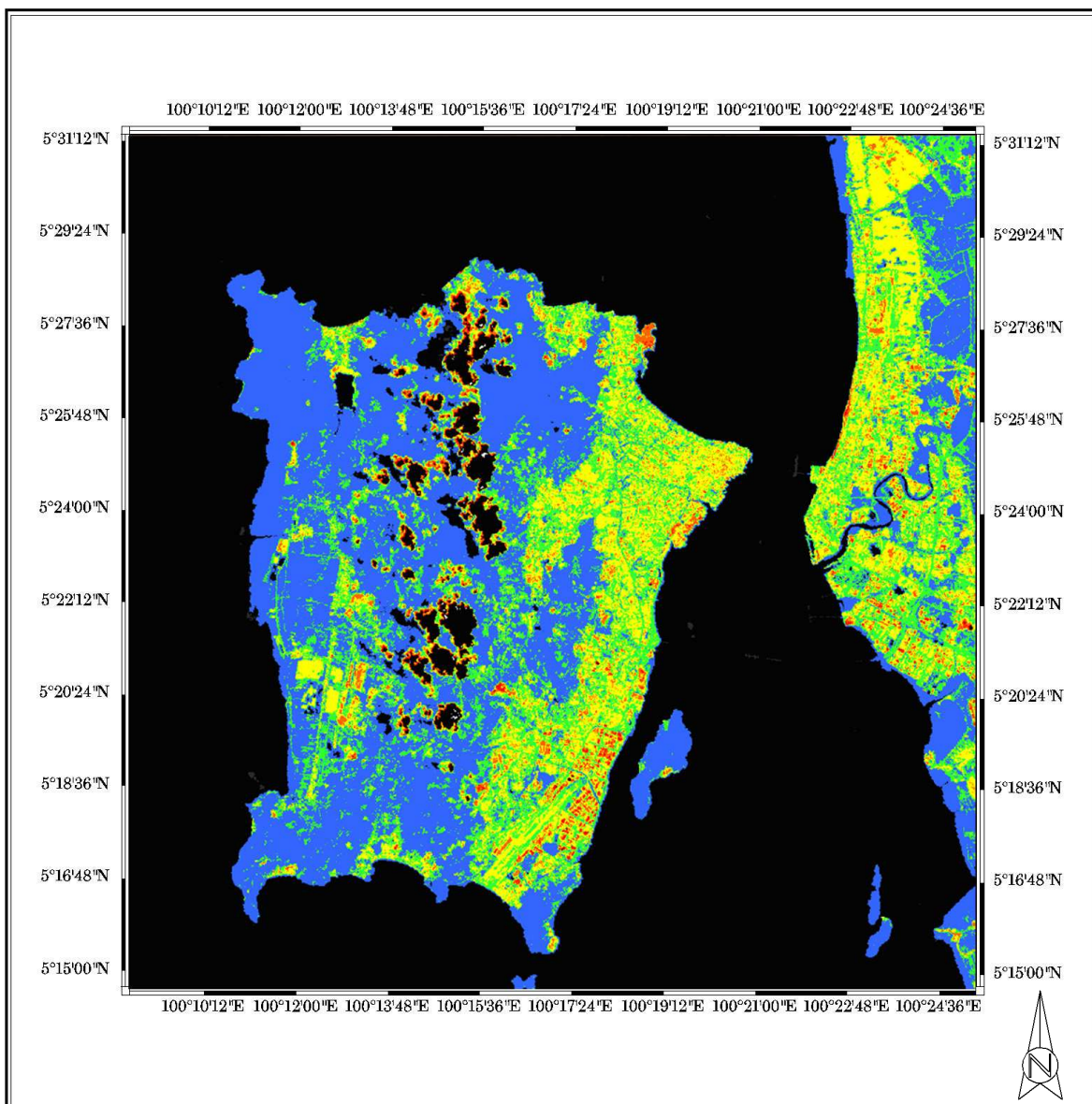


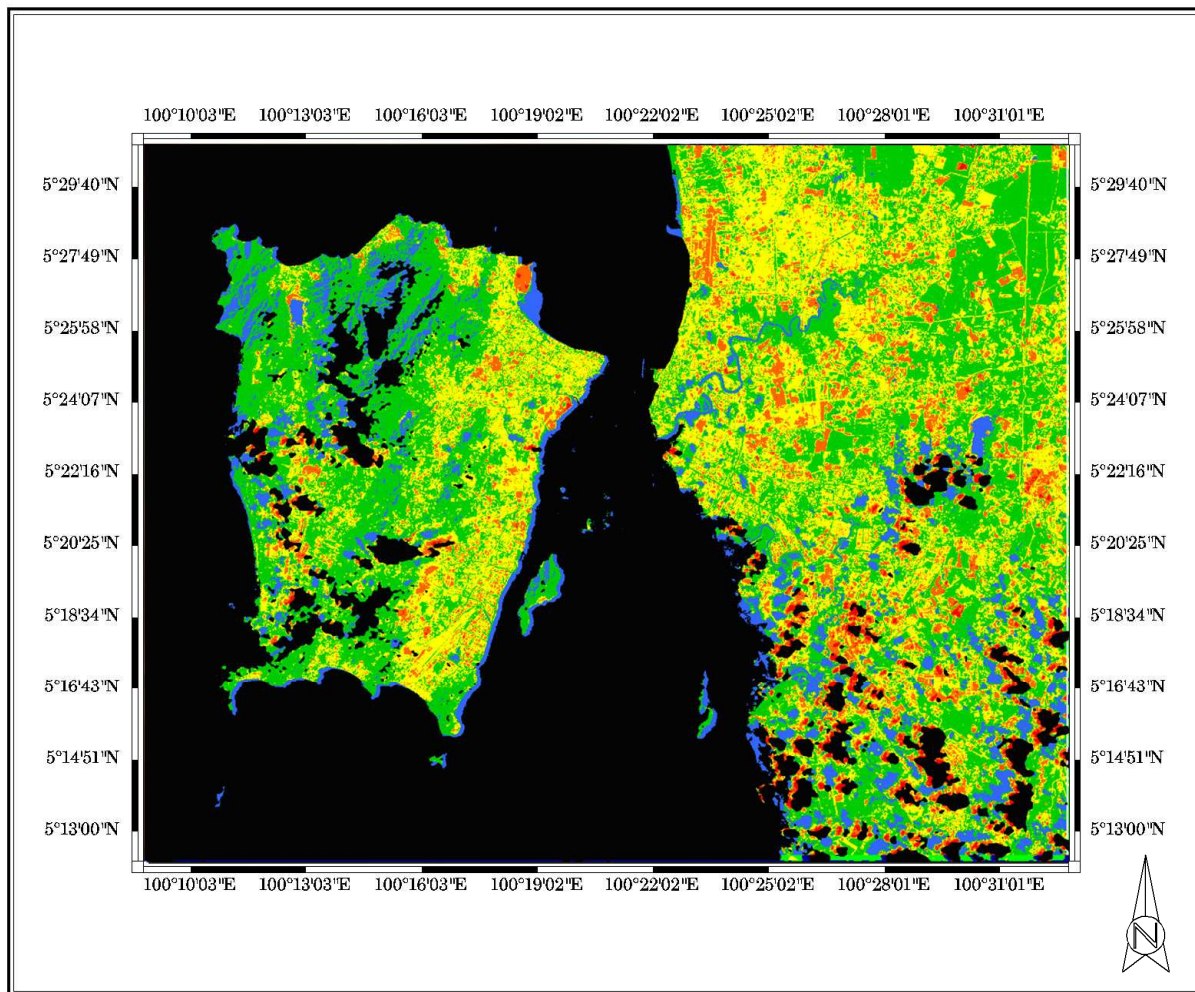
Fig. 14. Map of PM10 around Penang Island, Malaysia-5/2/2003



Legend

	<math><40 \mu\text{g}/\text{m}^3</math>		120-160 $\mu\text{g}/\text{m}^3$
	40-80 $\mu\text{g}/\text{m}^3$		>160 $\mu\text{g}/\text{m}^3$
	80-120 $\mu\text{g}/\text{m}^3$		Air dan Kawasan Awan

Fig. 15. Map of PM10 around Penang Island, Malaysia-19/3/2004



Legend



Fig. 16. Map of PM10 around Penang Island, Malaysia-2/2/2005

5. Conclusion

Image acquired from the satellite Landsat TM was successfully used for PM10 mapping over Penang Island, Malaysia. The developed algorithm produced a high correlation between the measured and estimated PM10 concentration. Further study will be carried out to verify the results. A multi regression algorithm will be developed and used in the

analysis. This study had shown the feasibility of using Landsat TM imagery for air quality study.

6. Acknowledgements

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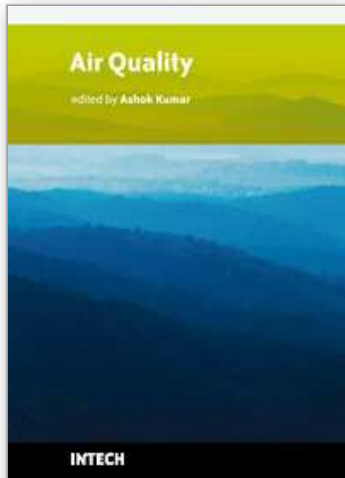
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Air pollution is about five decades or so old field and continues to be a global concern. Therefore, the governments around the world are involved in managing air quality in their countries for the welfare of their citizens. The management of air pollution involves understanding air pollution sources, monitoring of contaminants, modeling air quality, performing laboratory experiments, the use of satellite images for quantifying air quality levels, indoor air pollution, and elimination of contaminants through control. Research activities are being performed on every aspect of air pollution throughout the world, in order to respond to public concerns. The book is grouped in five different sections. Some topics are more detailed than others. The readers should be aware that multi-authored books have difficulty maintaining consistency. A reader will find, however, that each chapter is intellectually stimulating. Our goal was to provide current information and present a reasonable analysis of air quality data compiled by knowledgeable professionals in the field of air pollution.

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