# **RESEARCH ARTICLE**

### A time geographic approach for delineating areas of sustained wildlife use

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# 1 Abstract

2 Geographic information systems (GIS) are widely used for mapping wildlife movement patterns, and observed wildlife locations are surrogates for inferring on wildlife 3 4 movement and habitat selection. We present a new approach to mapping areas where 5 wildlife exhibit sustained use, which we term slow movement areas (SMAs). Nested 6 within the habitat selection concepts of home range and core areas, SMAs are an 7 additional approach to identifying areas important for wildlife. Our method for 8 delineating SMAs is demonstrated on a grizzly bear (Ursus arctos) case study examining 9 road density. Our results showed that subadult females had significantly higher road 10 densities within SMAs than in their PPA home ranges. The lowest road density was 11 found in the SMAs of adult male grizzly bears. Given increased mortality risks associated 12 with roads, female encampment near roads may have negative conservation implications. 13 The methods presented in this manuscript compliment recent developments to identify 14 movement suspension and intensively exploited areas defined from wildlife telemetry 15 data. SMA delineation is sensitive to missing data and best applied to telemetry data 16 collected with a consistent resolution.

Keywords: time geography; stopover ecology; GPS telemetry; potential path area;
grizzly bear

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## 20 1. Introduction

21 Due to improved GPS technology there has been an increase in availability of telemetry 22 data that has led to growth in movement analysis methods development (e.g., Thériault et 23 al. 1999, Dodge et al. 2008, Long and Nelson 2013). At the frontier of movement 24 research are wildlife studies that use movement as a surrogate for understanding 25 behaviour. Taking a classic spatial statistics perspective, the spatial pattern of observed 26 wildlife locations is an expression of spatial processes that are difficult to measure 27 directly (Nelson and Boots 2008). In this case, the spatial processes are biological and 28 originate from dynamic wildlife behaviour. Given that we cannot observe behaviour 29 continuously in space and time, patterns of movement are a surrogate measure for 30 behavioural states (Morales et al. 2004). For instance, Hunter (2007) determined that 31 foraging behaviours occurred when grizzly bears (Ursus arctos) were moving at a 32 velocity of less than 52m/minute. Food searching was associated with movement 33 velocities of 52 m/minute to 223 m/minute and active walking occurred at velocities of 34 greater than 223 m/minute.

35 Related to wildlife movement research is the use of telemetry data for 36 understanding spatial and temporal patterns of habitat selection (e.g., Berland *et al.* 2008) 37 Telemetry data represent discrete locations of an individual animal in space and time and 38 have been used extensively to study habitat selection by wildlife (Smulders et al. 2010). 39 Most habitat selection research employs the concept of home range or core area 40 (Smulders *et al.* 2012). A home range is typically defined as the area to which an animal 41 confines its normal movements (Burt 1943) and the core area is an intensively utilised 42 subset of the home range (Samuel and Green 1988). Recently, other concepts such as

intensively selected areas are also being employed (Benhamou and Roitte-Lambert 2012).
There are many ways to define a home range and core area, but arguably the most
common is by applying kernel density estimation to telemetry data to generate an
utilisation distribution. The 95% contour of the utilisation distribution is associated with
the home range while the 50% contour is associated with the core (Worton 1987). Spatial
units, like home range and core area, represented as discrete polygons, are integrated with
spatially continuous data on the physical environment to characterise habitat conditions.
The end goal is often to characterise the environmental elements (e.g., land cover types)
that can be best managed for the purpose of wildlife conservation (Bourbonnais <i>et al.</i>

52 2013).

Even with the availability of detailed wildlife location data and more continuous and timely landscape data from remotely sensed imagery, it has proven difficult to quantify links between movement, behaviour, and habitat. As typical with spatial pattern analysis, assumptions are required to make a linkage between spatial patterns of wildlife occurrence and processes of wildlife behaviour (Getis and Boots 1978). These assumptions pose difficulty due to subjective thresholds that are applied to patterns to categorise behaviour (Hunter 2007). Concepts of home range and core area have persisted in the literature, in part, because they are conceptually tidy and do not require inference on behaviour. Using home ranges and related concepts, habitats are associated with locations where wildlife are observed and the more often they are in observed in a location the greater the resource utilisation. Though home range and core areas are important for identifying habitats selected by wildlife, they cannot be employed to identify wildlife "use" areas, as the activity of wildlife at any given location is not known. 

# Annals of GIS

66	In this paper we revisit the notion of spatial-temporal patterns of resource
67	utilisation, selection, and use by defining slow movement areas (SMAs). SMAs are
68	defined at the finest spatial and temporal scale afforded by given telemetry data sets. We
69	argue that when animals are moving the slowest there is a high likelihood that the
70	selected or available habitat is being actively utilised for a specific biological function.
71	We do not classify use behaviour in SMAs, however likely behaviours are resting and
72	feeding or stopovers, depending on the scale of data.
73	We propose a new method for delineating SMAs by modifying an existing
74	technique for quantifying animal home ranges: the potential path area (PPA) home range
75	(Long and Nelson 2012). This method builds upon an existing analytical framework,
76	termed time geography (Hägerstrand 1970), useful for quantifying and examining the role
77	of spatial-temporal constraints on movement (Baer and Butler 1999). The PPA method
78	takes a pragmatic approach to movement analysis, focused on defining areas that are
79	spatially and temporally accessible. Given the importance of relating movement to habitat
80	conditions, the PPA polygons provide a simple approach for characterising habitats
81	associated with locations likely utilised. The benefit of using the existing PPA to define
82	new SMAs is demonstrated through a case study on grizzly bears in Alberta, Canada. We
83	show how within PPA home ranges, which represent habitat selection, and SMAs, which
84	represent likely habitat use, the density of roads varies and trends for males and females
85	are also opposite. Given that bears in Alberta are most likely to die near roads (Benn and
86	Herrero 2002), intensive habitat use near roads may negatively impact survival
87	(Bourbonnais et al. 2013).

88 2. Slow Movement Areas (SMAs)

89	While there are several methods for computing a home range, we based our approach for
90	computing SMAs on Long and Nelson (2012) and calculated a grizzly bear home range
91	using the PPA home range method. The PPA home range delineates the area accessible to
92	the animal given its sequence of telemetry fixes and a movement parameter ( $v_{max}$ –
93	defined as an animal's maximum travelling speed). The spatial area of accessibility
94	between two fixes can be defined based on the single parameter – vmax, and the time
95	difference between the two fixes, and is easily computed as a perfect ellipse shape.
96	(Figure 1 – upper panel). The PPA ellipse encompasses the entire area the animal could
97	have traversed based on its maximum travel speed and the location and time duration
98	between consecutive fixes. By combining the $n$ -1 ellipses, from a dataset of $n$ telemetry
99	fixes, the PPA home range is delineated (Long and Nelson 2012).
100	In order to compute SMAs we first define a statistic $m_i$ ( $i = 1n-1$ ) representing
101	the number of consecutive telemetry fixes that fall within each PPA ellipse (Figure $1 -$
102	lower panel). By definition, each ellipse will include, at a minimum, two fixes ( $m_i \ge 2$ ;
103	i.e., the current fix, and the next telemetry fix). The $m_i$ with the highest scores can then be
104	used to represent SMAs on a map, taking the highest score(s), or based on some
105	threshold. Mapping of SMAs involves taking the union of the $m_j$ -1 PPA ellipses of the $m_j$
106	fixes beginning with index <i>j</i> , where an SMA can be defined as:
107	$SMA = \bigcup PPA_{j(j+mj-1)})$

108 for each high scoring  $m_i$ . Spatially, the SMAs are sub-regions of the individual's home 109 range and represent the local accessibility space while encamped or slow moving. Once 110 delineated, SMA polygons can be treated much like home range polygons for analysing 111 underlying environmental characteristics. Similarly, because of how they are defined, the SMAs also represent a temporal sub-interval of the telemetry dataset, and this temporal
information can be used to further assess the timing of encamped and slow movement
behaviour.

115 The calculation of the PPA home range and SMAs will be impacted by the 116 selection of the  $v_{max}$  parameter. Parameterising  $v_{max}$  is subject to similar issues as have 117 been discussed for the bandwidth selection when using kernel density estimation to 118 define home ranges (Seamann et al. 1999, Gitzen et al. 2006, Nelson and Boots 2008). 119 Higher values of  $v_{max}$  will lead to the delineation of larger PPA home range and SMAs. 120 As  $v_{max}$  increases the animal is represented as being able to move more quickly and 121 therefore has more accessible habitat. Like kernel density bandwidth selection, selecting 122  $v_{max}$  will always be prone to some subjectivity (see Nelson and Boots 2008 for discussion 123 of bandwidth selection). We suggest that analysts use multiple confirmatory sources 124 when determining the most appropriate  $v_{max}$  parameter. Biological information on 125 maximum or typical speeds of travel can be compared to estimates generated from 126 observed data to build confidence in the  $v_{max}$  value selected.

127 The spatial-temporal extent and resolution of telemetry data will also impact the 128 interpretation, and indeed appropriateness, of home range and SMAs defined using PPA 129 approaches (Figure 2). Ideally, PPA approaches are applied when the spatial-temporal 130 resolutions and extents of telemetry fixes are similar throughout a dataset (see Wiens 131 1989 for discussion of scale). For instance, if wildlife data are collected every 20 minutes 132 in a 10 by 10 km area the data are relatively fine and SMAs are likely representation of 133 sleeping or feeding. In contrast, landscape scale trends, such as migratory stop-over 134 locations, could be identified when the SMA is defined for data collected once a day over

135 a broad area. Interpreting a SMA will be problematic if the resolution of the data is coarse 136 and the study area fine, as the areas delineated as SMA will be overgeneralised. 137 Similarly, if the study area is very large and the resolution of telemetry data fine, the 138 SMAs defined will likely be too small. Partitioning the telemetry data into smaller subsets prior to analysis may lead to more meaningful results. 139 140 Users should also be cautioned against defining SMAs in datasets that have 141 variable spatial-temporal resolutions and missing fixes. If some fixes are taken at both 142 one and four hour intervals, the longer intervals will have larger PPA ellipses and be 143 biased towards higher counts of consecutive points within the ellipse. When data are 144 sampled at varying resolutions the data can be partitioned by resolution, for separate 145 analysis and SMA delineation, or all the data downgraded to the coarsest resolution. Missing fixes are also problematic. If dropped fixes are not accounted for, PPA 146 147 ellipses could be artificially large and/or counts of consecutive points within the ellipse 148 low. If many fixes are dropped we recommend excluding that portion of the telemetry 149 data from SMA calculations to preserve analysis integrity. However, if only a few fixes 150 are missing it may be possible to clean the data by interpolating fixes. Given the 151 sensitivity of many methods to missing data (Frair et al. 2010) corrections, such as linear 152 interpolation based on curvilinear interpolation, have been demonstrated to improve data 153 quality (Tremblay et al. 2006). Another technical issue will arise when fixes are dropped 154 if an individual is denning or resting in an area that has poor signal coverage. No method 155 can pick up habitat selection or use in locations that for reasons of terrain or vegetation do not record telemetry fixes and many movement metrics are sensitive to missing data 156 157 (Laube and Purves 2011). However, since SMAs are intended to pick up slow movement,

158 which may be associated with resting, missed resting locations is an important omission 159 to consider. There is no systematic way to identify omitted resting area due to missing 160 data. Rather, when fix frequency becomes low or is missed for an extended period we 161 recommend manual assessment. 162 3. Case study 163 Our methodology to define SMAs is demonstrated with a case study of grizzly bears in 164 Alberta, Canada. In the Kakwa region of west-central Alberta, grizzly bears share their 165 habitat with many anthropogenic disturbances that are affecting the bears' traditional use 166 and selection of habitat. To illustrate the utility of our methodology, we examined the 167 density of roads within the SMAs compared to the PPA home range. Road density has 168 been found to correlate with mortality risk and reduced survival in grizzly bears (Benn 169 and Herrero 2002, Nielsen et al. 2004), yet areas with high road densities are often 170 selected as habitat (Roever et al. 2008, Graham et al. 2010, Stewart et al. 2013). It is not 171 fully understood why bears appear to select habitat with roads, but it has been speculated 172 that roadside areas offer bear food (Roever et al. 2008, Graham et al. 2010, Stewart et al. 2013). 173 174 Differences in habitat selection by male and female grizzly bears are becoming 175 increasingly documented as the body of grizzly bear research grows. The much larger males are known to have larger home ranges (Proctor et al. 2004, Roever et al. 2008, 176

177 Graham *et al.* 2010) and greater daily movement rates when compared to females

178 (Boulanger *et al.* 2013). In contrast, females have been found to select habitat containing

179 roads more than males (Roever *et al.* 2008, Graham *et al.* 2010, Stewart *et al.* 2013).

180 Female selection of roads is of concern, especially for a threatened population, given that

female survival is paramount for population viability (Eberhardt *et al.* 1994, Stewart *et al.*2013). Understanding the behaviour of female grizzly bears associated with roads will
provide important conservation information for those tasked with land-use decision
making.

185 3.1 Study area and data

The study area for this research is an 8308 km<sup>2</sup> landscape in the Kakwa region of west-186 187 central Alberta, Canada. The elevation ranges from 549 m to 2446 m, and the area 188 comprises a diverse and multi-use landscape. Resource extraction industries have been 189 active in the area for a number of decades (White *et al.* 2011), with most disturbances in 190 the area arising from the forest industry and oil and gas exploration (Schneider 2002). 191 A dataset of GPS locations collected over 2005–2010 from 25 grizzly bears in the 192 study area were provided by the Foothills Research Institute Grizzly Bear Program 193 (Hinton, AB). The FRI researchers followed the accepted protocols of the Canadian 194 council of animal care for the safe handling of bears (animal use protocol number 195 20010016) (Stenhouse and Munro 2000). Bears were fitted with Televilt/Followit brand 196 GPS collars (Lindesburg, Sweden). We obtained road data for the study area from

197 Alberta Environment and Sustainable Resource Development and updated it through

heads up digitising of medium and high resolution satellite imagery (SPOT and airphotos).

200 3.2 Analysis

201 By definition, SMAs are delineated using both the spatial and the temporal structure of

202 telemetry data. Thus, we began our analysis by correcting for missing fixes that are

203 inevitable with GPS-based telemetry systems (Rempel et al. 1995). An interpolation

204	algorithm was used to accommodate for missing fixes in order to generate trajectories
205	with consistent sampling intervals (regular trajectories—Calenge et al. 2009). When the
206	GPS-signal was disrupted for an extended period of time (i.e., > 4 fixes) we analysed the
207	bear trajectory separately on either side of the disruption.
208	For each individual bear we calculated the PPA home range and SMA (Figure 3).
209	We took a simple approach to SMA analysis here using only the longest encamped period
210	(i.e., the $m_i$ with the highest score) to generate the SMA. Road density within individual
211	bear home ranges and SMAs were then calculated and summarised by age and sex. We
212	excluded the SMA from the home range when calculating road density to compare
213	between the home range and SMA. Grizzly bears less than five years old are considered
214	subadults and their selection of habitat has been shown to be different from adult bears
215	(Mueller et al. 2004). We partitioned bears by age (subadult or adult) and by sex. For
216	each age-sex class (adult females, adult males, subadult females, subadult males) we
217	assessed the statistical differences in the road density within home ranges and SMAs by
218	comparing frequency distributions using a Mann-Whitney U statistical test.
219	3.3 Results
220	For adult females, the average PPA home range was 466.83 km <sup>2</sup> , whereas the average
221	SMA size was 117.24 km <sup>2</sup> . The average PPA home range for sudadult females was
222	540.00 km <sup>2</sup> and the average SMA was calculated to be 153.91 km <sup>2</sup> . When all female data
223	were combined, the average PPA home range was 479.44 km <sup>2</sup> and the average SMA was
224	131.72 km <sup>2</sup> .

Adult male grizzly bears were found to have an average PPA home range that was 674.58 km<sup>2</sup> and an average SMA that was 149.21 km<sup>2</sup>. The average PPA home range for

227	subadult males was 560.70 $\text{km}^2$ whereas their average SMA was found to be 136.81 $\text{km}^2$
228	When all males were considered together, the average PPA home range was $651.49 \text{ km}^2$
229	and the SMA was calculated to be $144.53 \text{ km}^2$ .
230	Average road density was calculated for all groups in both the PPA home range
231	and the SMAs (Table 1, Figure 4 and Figure 5). Road density in the SMAs for adult
232	females was very similar to the road density in the PPA home range (0.60 km/km <sup>2</sup> and
233	0.59 km/km <sup>2</sup> , respectively) (Table 1 and Figure 5). However, for subadult females, the
234	road density was significantly higher ( $p = 0.0209$ ) in the SMA compared to the PPA
235	home range (0.66 km/km <sup>2</sup> and 0.50 km/km <sup>2</sup> , respectively)(Table 1 and Figure 5).
236	In general, male grizzly bears were found to have lower road densities in their
237	SMAs compared to their PPA home ranges (Table 1 and Figure 4). The lowest road
238	density was found in the SMAs of adult males (0.43 km/km <sup>2</sup> ). Males also generally had
239	lower road densities in both their SMAs and PPA home ranges (0.46 $\text{km/km}^2$ and 0.52
240	km/km <sup>2</sup> , respectively) compared to their female counterparts (0.63 km/km <sup>2</sup> and 0.57

241 km/km<sup>2</sup>, respectively).

## 242 **4. Discussion**

Grizzly bears often rest adjacent to sites recently used for feeding (Phillips 1987) and it is reasonable to assume that the low mobility activities in the SMAs consisted primarily of feeding/foraging and resting. Previous research has demonstrated the selection of roads by grizzly bears (Chruszcz *et al.* 2003, Graham *et al.* 2010, Stewart *et al.* 2013), yet were not able to provide movement details. Roads have been associated with increased mortality in grizzly bears (Benn and Herrero 2002) and it is important to fully recognise their attraction to bears when making land-use decisions that support conservation (see

### Annals of GIS

Stewart et al. 2013 for a more in-depth discussion) and aid in population recovery efforts. 250 251 It is concerning that the subadult females in our study had a significantly greater 252 concentration of roads in their SMAs compared to the remainder of their home ranges. 253 The survival of vulnerable subadult females into the adult breeding stage is essential for 254 population viability (Mueller et al. 2004). While previous studies have observed the 255 selection of roads by subadult females (Mueller et al. 2004), the results of our case study 256 provide insights into the behaviour associated with roads. 257 It is interesting that the male grizzly bears had fewer roads in their SMAs 258 compared to the remainder of their home range and also to female bears. A previous 259 study using the same database had found male grizzly bears to select natural edge habitats 260 over anthropogenic edges (Stewart et al. 2013). Our case study has enabled us to 261 determine that slow movement behaviours in males are associated with areas with fewer roads. A road density of 0.6 km/km<sup>2</sup> has been previously postulated as the limit for 262 263 naturally functioning landscapes containing sustained populations of large predators 264 including grizzly bears (Forman and Alexander 1998). Our study suggests that areas with 265 lower road densities appear to be most desirable for adult males' encampment and it is 266 possible that the female bears are being competitively excluded from these areas by more 267 dominant conspecifics (Mattson et al. 1987; Edwards et al. 2011). 268 Methods for analysing spatial-temporal data have been touted as an opportunity 269 area for spatial science development (Nelson 2012). Movement data are inherently spatial 270 and temporal and there are many examples of recent developments in methods for 271 quantifying movement in people (Jankowski et al. 2010), wildlife (Langrock 2012), and 272 traffic (Andrienko and Andrienko 2013). SMA delineation compliments recent progress

in movement science, such as the development of methods to identify suspension in
human movement (Orellana and Wachowicz 2011), stopover ecology (Sawyer and
Kauffman 2011), and areas intensively exploited by wildlife (Benhamou and RiotteLambert 2012).

277 Wildlife researchers require a range of methods to characterise different types of 278 movement patterns. The most simple movement pattern measure is velocity obtained by 279 dividing the spatial distance by the time difference of two consecutive fixes (Brillinger et 280 al. 2004, Chapman et al. 2007). Velocity (speed, time lag, or step length) (Brillinger et al. 281 2004, Calenge et al. 2009) and other metrics (turning angle and bearing) (Turchin 1986, 282 Calenge et al. 2009) are related to behaviour by defining arbitrary thresholds. A one-to-283 one relationship between spatial patterns of movement and behaviour is difficult to define 284 making it problematic to relate behaviour and habitat. As well, it is often desirable to 285 associate movement behaviour with an area, as is often the case in habitat selection 286 studies, point based representations are limited. Another potential limitation of basic 287 velocity measures is that they are computed based on only two consecutive fixes, 288 ignoring potentially useful information from larger consecutive intervals within the 289 telemetry data. Turning angle is typically computed on three points but has similar 290 limitations.

The theory of SMAs links conceptually with existing notions of home range and core area delineation (Worton 1987). Without requiring a classification of wildlife behaviour, SMAs allow us to define areas that have a high probability of resource use. The nature of that resources use will vary depending on species and scales of data. Like home range and core area delineation the strength of SMAs lies in the assumption that

296	spatial patterns are expressions of spatial processes (Getis and Boots 1978). A single
297	spatial pattern can be related to many different processes making behaviour difficult to
298	infer. Therefore, methods that can identify utilisation, without requiring one-to-one
299	relationships with behaviour, are important for wildlife research and support assessment
300	of utilised habitats and wildlife conservation. However, a unique component of both the
301	PPA and SMA methods is the utilisation of the temporal component of the data. The
302	selection of SMAs is consecutive in time. With increasingly available high resolution
303	telemetry data, the SMA approach to identifying habitats associated with sub-regions of
304	the home range associated with encamped or slow movement behaviours. In the case of
305	the grizzly bear, identifying SMAs may indicate critical foraging regions that are
306	important in conservation management efforts since grizzly bears require almost
307	continuous feeding to meet their nutritional needs (Rode et al. 2001).
308	As SMAs are an extension of the PPA home range approach there are also strong
309	ties to other recently developed path-based measures of animal home range, namely the
310	Brownian bridge home range (Bullard 1999, Horne et al. 2007), and time geographic
311	kernel density estimation (Downs 2010, Downs et al. 2011). The PPA home range
312	represents a direct measure of spatial range (as a spatial polygon) while both the
313	Brownian bridge and time geographic kernel density estimation methods first estimate a
314	utilisation distribution, followed by extracting a home range polygon. The definition of
315	SMAs within a PPA home range is another mechanism for understanding utilisation,
316	though it is a measure of encampment rather than percentage time spent at a location. A
317	benefit of the SMA delineation is that areas are defined using maximum speed as the only
318	subjective parameter. Our approach specifically does not identify frequently revisited

areas and alternative approaches (e.g., Benhamou and Riotte-Lambert 2012) are more
appropriately designed specifically for discovering these revisitation areas, for example
associated with important movement corridors.

322 To calculate SMAs, every telemetry location is assigned both an ellipse and a 323 value  $(m_i)$  identifying the number of consecutive points that fall within that ellipse. As 324 such, it is possible to map how long an animal was in an area for all PPA ellipses or 325 telemetry locations. Different lengths of utilisation could be linked with different 326 behaviours (i.e., foraging, resting, and travelling); however, as with velocities, linkages to 327 behaviour require that subjective thresholds be defined. In this analysis we defined the 328 SMA using only the largest  $-\max(m_i)$  – value, identifying a SMA, however in many 329 applications it will be advantageous incorporate, for example, the 10 largest values of  $m_i$ . 330 This may be especially important with larger telemetry datasets covering long temporal 331 durations, where multiple SMAs could identify recurring behaviour associated with 332 sustained use and low mobility rates.

333 Future research could develop techniques for more objective definitions of 334 movement pattern thresholds. For instance, using theory from spatial statistics it may be 335 possible to begin teasing apart when various movement patterns are most likely 336 realisations of different processes (Getis and Boots 1978, Smulders et al. 2010). Similar 337 to variability in the nature of home ranges, how the SMA is utilised is related to the 338 spatial and temporal scales of the telemetry data. As in the grizzly bear example, when 339 data are hourly or finer, the behaviours most likely associated with SMAs are feeding or 340 resting. If telemetry data are collected at coarse temporal resolutions and extents the 341 SMAs will reflect broader scale processes such as migratory stopovers.

## **5.** Conclusions

343 New methods for characterising wildlife movement patterns will give researchers greater 344 flexibility in the types of hypotheses investigated. We present a new approach to 345 delineating areas where an animal exhibits sustained use. Similar to home ranges and 346 core areas, SMAs are areas where spatial patterns indicate habitat selection and do not 347 require explicit categorisation of behaviour. However, SMAs are related to encampment 348 and will represent a range of short-term behaviours such as foraging or resting when 349 telemetry data are collected frequently, and migratory stopover locations for data sets 350 with a long temporal extent. Regardless of scale, the areas defined by the SMA have a 351 high likelihood of wildlife resource use. SMA delineation methods require consistent 352 spatial-temporal resolutions and minimal missing data. Future research should investigate 353 how a time geographic framework, such as the PPA ellipses presented here, can be used 354 to map a range of habitat utilisation behaviours based on length of time spent in each 355 area.

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### 364 **References**

365	Andrienko, N., and Andrienko, G., 2013. Visual analytics of movement: An overview of
366	methods, tools and procedures. Information Visualization, 12(1), 3-24.
367	Berland, A., Nelson, T., Stenhouse, G., Graham, K., & Cranston, J. (2008). The impact of
368	landscape disturbance on grizzly bear habitat use in the Foothills Model Forest,
369	Alberta, Canada. Forest ecology and management, 256(11), 1875–1883.
370	Benhamou, S. and Riotte-Lambert, L., 2012. Beyond the Utilisation Distribution:
371	identifying home range areas that are intensively exploited or repeatedly visited.
372	Ecological Modelling, 227, 112–116.
373	Benn, B. and Herrero, S., 2002. Grizzly bear mortality and human access in Banff and
374	Yoho National Parks, 1971–98. Ursus, 13, 213–221.
375	Boulanger, J., Cattet, M., Nielsen, S.E., Stenhouse, G., and Cranston, J., 2013. Use of
376	multi-state models to explore relationships between changes in body condition,
377	habitat and survival of grizzly bears Ursus arctos horribilis. Wildlife Biology,
378	19(3), 274–288.
379	Bourbonnais, M.L., Nelson, T.A., Cattet, M.R., Darimont, C.T., and Stenhouse, G.B.,
380	2013. Spatial analysis of factors influencing long-term stress in the grizzly bear
381	(Ursus arctos) population of Alberta, Canada. PloS one, 8(12), e83768.
382	Brillinger, D.R., Preisler, H.K., Ager, A.A. and Kie, J.G., 2004. An exploratory data
383	analysis (EDA) of the paths of moving animals, Journal of Statistical Planning
384	and Inference, 122(1-2), 43–63.
385	Bullard, F., 1999. Estimating the home range of an animal: A Brownian bridge approach.
386	Master of Science, University of North Carolina, Chapel Hill.

387	Burt, W. H., 1943. Territoriality and home range concepts as applied to mammals.
388	Journal of Mammalogy, 24, 346-352.
389	Calenge, C., Dray, S. and Royer-Carenzi, M., 2009. The concept of animals' trajectories
390	from a data analysis perspective. Ecological Informatics, 4, 34-41.
391	Chapman, D.S., Dytham, C. and Oxford, G.S., 2007. Landscape and fine-scale
392	movements of a leaf beetle: the importance of boundary behaviour. Oecologia,
393	154(1), 55–64.
394	Chruszcz, B. Clevenger, A.P., Gunson, K.E. and Gibeau, M.L., 2003. Relationships
395	among grizzly bears, highways, and habitat in the Banff-Bow Valley, Alberta,
396	Canada. Canadian Journal of Zoology, 81, 1378–1391.
397	Dodge, S., Weibel, R. and Lautenschütz, A.K., 2008. Towards a taxonomy of movement
398	patterns. Information Visualization, 7(3-4), 240–252.
399	Downs, J.A., 2010. Time-geographic density estimation for moving point objects.
400	Lecture Notes in Computer Science, 6292, 16–26.
401	Downs, J.A., Horner, M.W. and Tucker, A.D., 2011. Time-geographic density estimation

- 402 for home range analysis. *Annals of GIS*, 17, 163–171.
- 403 Eberhardt L.L., Blanchard, B.M. and Knight, R.R., 1994. Population trend of the

404 Yellowstone grizzly bear as estimated from reproductive and survival rates.

- 405 *Canadian Journal of Zoology*, 72, 360–363
- 406 Edwards, M.A., Derocher, A.E., Hobson, K.A., Branigan, M., and Nagy, J.A., 2011. Fast
- 407 carnivores and slow herbivores: differential foraging strategies among grizzly
- 408 bears in the Canadian Arctic. *Oecologia*, 165, 877–889.

409	Forman, R.T.T., and Alexander, L.E., 1998. Roads and their major ecological effects.
410	Annual Review of Ecology and Systematics, 29, 207–231.
411	Frair, J.L., Fieberg, J., Hebblewhite, M., Cagnacci, F., DeCesare, N.J., and Pedrotti, L.,
412	2010. Resolving issues of imprecise and habitat-biased locations in ecological
413	analyses using GPS telemetry data. Philosophical Transactions of the Royal
414	Society B: Biological Sciences 365.1550 (2010): 2187–2200.
415	Getis, A. and Boots, B., 1978. Models of Spatial Processes. Cambridge: Cambridge
416	University Press.
417	Gitzen, R.A., Millspaugh, J.J. and Kernohan, B.J., 2006. Bandwidth selection for fixed-
418	kernel analysis of animal utilization distributions. Journal of Wildlife
419	Management, 70, 1334–1344.
420	Graham, K., Boulanger, J., Duval, J. and Stenhouse, G., 2010. Spatial and temporal use
421	of roads by grizzly bears in west-central Alberta. Ursus, 21, 43–56.
422	Hägerstrand, T., 1970. What about people in regional science? Papers of the Regional
423	Science Association, 24, 7–21.
424	Harris, S., Cresswell, W.J., Forde, P.G., Trewhella, W.J., Woollard, T., and Wray, S.,
425	1990. Home range analysis using radio tracking data- a review of problems and
426	techniques as applied to the study of mammals. Mammal Review, 20(2-3), 97-
427	123.
428	Horne, J.S., Garton, E.O., Krone, S.M. and Lewis, J.S., 2007. Analyzing animal
429	movements using brownian bridges. Ecology, 88, 2354–2363.
430	Hunter, A., 2007. Sensor-Based Animal Tracking. PhD Diss., Department of Geomatics

431 Engineering, University of Calgary.

432	Kie, J.G., Matthiopoulos, J., Fieberg, J., Powell, R.A., Cagnacci, F., Mitchell, M.S.,
433	Gaillard, JM. and Moorcroft, P.R., 2010. The home-range concept: are
434	traditional estimators still relevant with modern telemetry technology?
435	Philosophical Transactions of the Royal Society B, 365, 2221–2231.
436	Jankowski, P., Andrienko, N., Andrienko, G., and Kisilevich, S., 2010. Discovering
437	landmark preferences and movement patterns from photo postings. Transactions
438	in GIS, 14(6), 833–852.
439	Johnson, D.S., London, J.M., Lea, M.A., and Durban, J.W., 2008. Continuous-time
440	correlated random walk model for animal telemetry data. Ecology, 89(5), 1208-
441	1215.
442	Langrock, R., King, R., Matthiopoulos, J., Thomas, L., Fortin, D. and Morales, J.M.,
443	2012. Flexible and practical modeling of animal telemetry data: Hidden Markov
444	models and extensions. Ecology, 93(11), 2336–2342.
445	Laube, P. and Purves, R., 2011. How fast is a cow? Cross-scale analysis of movement
446	data. Transactions in GIS, 15, 401-418.
447	Laver, P.N. and Kelly, M. J. 2008. A critical review of home range studies. The Journal
448	of Wildlife Management, 290–298.
449	Long, J.A. and Nelson, T.A., 2012. Time geography and wildlife home range delineation.
450	Journal of Wildlife Management, 76(2), 407–413.
451	Long, J.A. and Nelson, T.A., 2013. A review of quantitative methods for movement data.
452	International Journal of Geographical Information Science, 27(2), 292–318.
453	Mattson, D.J., Knight, R.R., and Blanchard, B.M., 1987. The effects of developments and
454	primary roads on grizzly bear habitat use in Yellowstone National Park,

	22T.A. Nelson et al.	
455	Wyoming. International Conference of Bear Research and Manag	gement, 7, 259–
456	273.	
457	Morales, J.M., Haydon, D.T., Frair, J., Holsinger, K.E. and Fryxell, J.M.,	2004.
458	Extracting more out of relocation data: building movement model	s as mixtures of
459	random walks. <i>Ecology</i> , 85(9), 2436–2445.	
460	Mueller, C., Herrero, S., and Gibeau, M.L., 2004. Distribution of subadul	t grizzly bears
461	in relation to human development in the Bow River Watershed, A	lberta. Ursus,
462	15(1), 35–47.	
463	Nielsen, S.E., Herrero, S., Boyce, M.S., Benn, B., Mace, R.D., Gibeau, M	I.L. and Jevons,
464	S., 2004. Modelling the spatial distribution of human-caused grizz	ly bear
465	mortalities in the Central Rockies Ecosystem of Canada. Biologica	al Conservation,
466	120, 101–113.	
467	Orellana, D. and Wachowicz, M., 2011. Exploring patterns of movement	suspension in
468	pedestrian mobility. Geographical Analysis, 43, 241–260.	
469	Phillips, M.K., 1987. Behavior and habitat use of grizzly bears in northea	stern Alaska.
470	International Conference on Bear Research and Management, 7,	159–167.
471	Proctor, M.F., McLellan, B.N., Strobeck, C. and Barclay, R.M.R., 2004.	Gender-specific
472	dispersal distances of grizzly bears estimated by genetic analysis.	Canadian
473	Journal of Zoology, 82, 1108–1118.	
474	Rempel, R.S. Rodgers, A.R. and Abraham, K.F., 1995. Performance of a	GPS animal
475	location system under boreal forest canopy. Journal of Wildlife M	anagement, 59,
476	543–551.	

- 477 Rode, K.D., Robbins, C.T. and Shipley, L.A., 2001. Constraints on herbivory by grizzly
  478 bears. *Oecologia*, 128, 62–71.
- 479 Roever, C.L., Boyce, M.S. and Stenhouse, G.B., 2008. Grizzly bears and forestry II:
- 480 Grizzly bear habitat selection and conflicts with road placement. *Forest Ecology*481 *and Management*, 256: 1262–1269.
- 482 Samuel, M.D. and Green, R.E., 1988. A revised test procedure for identifying core areas
  483 within the home range. *Journal of Animal Ecology*, 57(3), 1067–1068.
- 484 Seaman, D.E., Millspaugh, J.J., Kernohan, B.J., Brundige, G.C., Raedeke, K.J. and
- 485 Gitzen, R.A., 1999. Effects of sample size on kernel home range estimates. *The*486 *Journal of Wildlife Management*, 739-747.
- 487 Sawyer, H. and Kauffman, M.J., 2011. Stopover ecology of a migratory ungulate.

488 *Journal of Animal Ecology*, 80, 1078–1087.

489 Schneider, R.R., 2002. Alternative Futures: Alberta's Boreal Forest at the Crossroads.

490 Edmonton: The Federation of Alberta Naturalists.

- 491 Smulders, M., Nelson, T.A., Jelinski, D.E., Nielsen, S.E., and Stenhouse, G.B., 2010. A
- 492 spatially explicit method for evaluating accuracy of species distribution models.
  493 *Diversity and Distributions*, 16(6), 996–1008.
- 494 Smulders, M., Nelson, T.A., Jelinski, D.E., Nielsen, S.E., Stenhouse, G.B., and Laberee,
- K., 2012. Quantifying spatial-temporal patterns in wildlife ranges using STAMP:
  A grizzly bear example. *Applied Geography*, *35*(1), 124–131.
- 497 Stenhouse, G.B. and Munro, R.H.M. 2000. Foothills Model Forest Grizzly Bear Research
- 498 *Program 1999 Annual Report.* Hinton: Foothills Model Forest.

499	Stewart, B.P., Nelson, T.A., Laberee, K., Nielsen, S.E., Wulder, M.A., and Stenhouse,
500	G.B., 2013. Quantifying grizzly bear selection of natural and anthropogenic
501	edges. Journal of Wildlife Management, doi: 10.1002/jwmg.535.
502	Thériault, M., Claramunt, C. and Villeneuve, P., 1999, A spatio-temporal taxonomy for
503	the representation of spatial set behaviours, in Spatio-temporal Database
504	Management, Springer-Verlag, LNCS 1678, pp. 1–19.
505	Tremblay, Y., Shaffer, S.A., Fowler, S.L., Kuhn, C.E., McDonald, B.I., Weise, M.J.,
506	Bost, CA., et al. (2006). Interpolation of animal tracking data in a fluid
507	environment. Journal of Experimental Biology, 209(1), 128-140.
508	Turchin, P., 1986. Modelling the effect of host patch size on Mexican bean beetle
509	emigration, <i>Ecology</i> , 67(1), 124–132.
510	White, J.C., Wulder, M.A., Gomez, C. and Stenhouse, G.B., 2011. A history of habitat
511	dynamics: characterising 35 years of stand replacing disturbance. Canadian
512	Journal of Remote Sensing, 37(2), 234–25.
513	Wiens, J.A. 1989. Spatial scaling in ecology. Functional ecology, 3, 385-397.
514	Worton, B.J., 1987. A review of models of home range for animal movement. Ecological
515	Modelling, 38(3), 277–298.
516	

	Road Density (km/km²)			P-value	N	
	in PPA HR		in SMA			
	Mean	CoV	Mean	CoV		
adult female	0.59	0.24	0.60	0.36	0.7401	53
subadult female	0.50	0.19	0.66	0.42	0.0209	15
adult male	0.53	0.43	0.43	0.70	0.8373	33
subadult male	0.60	0.28	0.50	0.63	0.2732	20
female	0.57	0.24	0.63	0.22	0 1503	68
Ternate	0.57	0.24	0.05	0.22	0.1505	00
male	0.52	0.39	0.46	0.30	0.4634	53

519 Table 1. A comparison between the road density in the SMAs and the HR. Statistical

520 significance was determined using a Mann Whitney U test. Subadult females were found

521 to have significantly different road density in their SMAs compared to their HRs. N

522 represents the sum of individual bears by season for each year of the study (2005-2010).

523





Figure 1. Delineating Potential Path Area (PPA) and calculating the Slow Movement Area (SMA). The upper panel shows how the space-time prism contains all sets of accessible locations given two telmetry fixes,  $t_1$  and  $t_2$ . By combining several PPA ellipses a PPA home range is defined (see Long and Nelson 2012). In the lower panel, all consecutive telemetry locations within a PPA ellipse are counted in the calculation of the SMA. The PPA ellipse containing the largest number of consecutive telemetry fixes is used as the basis for the SMA.



535 Figure 2. Data resolutions and extents most appropriate for use with PPA home range and

536 SMA delineation do not mix scale. Grey areas indicate appropriate combinations of data

537 resolutions and extents for applying SMA delineation. When the scales are mixed the

538 SMA defined will be overly general and likely too large (upper left) or so small relative

539 to the space-time extent that it is not useful.



542 Figure 3. Defining the PPA home range and SMA for one male bear.



546 Figure 4. Box plots comparing the road density within all female and all male bear PPA

547 home ranges to their SMAs.



550 Figure 5. Box plots of road densities within PPA home ranges compared to SMAs for

adult females, subadult females, adult males and subadult males.