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# **Foraging Seabirds Respond To An Intermittent Meteorological Event In A Coastal Environment**

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## ABSTRACT

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Temporal variations in the numbers of foraging seabirds usually coincide with concurrent variations in physical processes influencing prey availability. Responses to periodic tidal currents are commonly reported, with certain tidal states being favoured. By contrast, responses to intermittent meteorological events have rarely been reported, even though wind-driven exchanges of water masses or intrusion of estuarine plumes could have similar consequences. As large-scale offshore constructions (e.g. aquaculture, coastal defences, ports and marine renewable energy installations) and climate variations alter periodic tidal currents and intermittent meteorological events, respectively, quantifying responses to these physical processes identifies potential impacts on seabird communities. This study quantifies responses of foraging seabirds to physical processes in the Ria de Vigo, north-western Spain. The numbers of foraging European Shags *Phalacrocorax aristotellus* and Yellow-legged Gulls *Larus michahellis* showed no response to variations in current direction and speed. By contrast, both increased in number during an estuarine plume intrusion (the Western Iberian Buoyant Plume: WIBP) following an extreme river discharge event and southerly winds. These increases may be explained by the temporary combination of marine and brackishwater fauna, increasing prey biomass. The frequency of extreme river discharge events is likely to decrease in north-western Spain. If WIBP intrusions consistently enhance prey availability, observations of large numbers of foraging seabirds using the ria could become rarer.

**Key Words:** estuarine plume, foraging ecology, European Shag, *Larus michahellis*, *Phalacrocorax aristotellus*, vessel-based surveys, Yellow-legged Gull

## INTRODUCTION

For foraging seabirds, coastal environments represent important habitats due to physical processes that enhance prey availability (Cox *et al.* 2018). However, numerous physical processes in coastal environments are susceptible to anthropogenic-driven changes. Large-scale offshore constructions (e.g. aquaculture, coastal defences, ports and marine

35 renewable energy installations; Carter 2013) alter tidal currents (Cazenave *et al.* 2016, De  
36 Dominicis *et al.* 2017, Fraser *et al.* 2017, Shields *et al.* 2011) whereas climate change and  
37 oscillations (e.g. North Atlantic Oscillation, El Nino Southern Oscillation) alter  
38 meteorological events (Stenseth *et al.* 2003, Harley *et al.* 2006). Identifying the responses of  
39 foraging seabirds to tidal currents and meteorological events in coastal environments would  
40 highlight potential impacts of anthropogenic-driven changes.

41

42 In coastal environments, periodic changes in the direction/speed of tidal currents and depth  
43 across ebb-flood cycles are a conspicuous physical process (Simpson *et al.* 2012). These  
44 changes influence prey availability. For seabirds targeting pelagic prey, a certain current  
45 direction or speed could advect prey from productive neighbouring areas, increasing  
46 encounters with prey (Zamon 2001). In other cases, certain combinations of current  
47 direction/speed and topography create turbulent eddies and shear-lines, entraining and  
48 aggregating prey (Johnston *et al.* 2007). For seabirds targeting benthic prey, the energetic  
49 cost of dives is reduced at slow current speeds and shallow depths, increasing the  
50 accessibility of prey (Heath *et al.* 2010). Studies showing the number of foraging seabirds  
51 increasing during certain tidal states are numerous and widespread (Hunt *et al.* 1999,  
52 Benjamins *et al.* 2015; Waggitt *et al.* 2016a, 2016b).

53

54 In some coastal environments however, meteorological events (e.g. extreme river discharge  
55 or intense wind) also represent important physical processes. Estuarine plumes following  
56 extreme river discharge events alter salinity and temperature (Gillanders *et al.* 2002), whereas  
57 exchanges of water masses during intense wind events have similar effects (Kämpf *et al.*  
58 2016). As with tidal currents, these meteorological events could also influence prey  
59 availability. For instance, onshore advection of productive water masses encourage prey to  
60 form denser schools (Benoit-Bird *et al.* 2019). Estuarine plumes encourage brackish-water  
61 species into the open-ocean, increasing prey biomass (Kingsford *et al.* 1994). The frequency  
62 of these meteorological events is usually seasonal, with the highest numbers of foraging  
63 seabirds seen when favourable meteorological events are most likely (Cox *et al.* 2018).  
64 However, the timing of individual meteorological events within seasons are intermittent and  
65 unpredictable. Studies showing changes in the number of foraging seabirds during an  
66 intermittent meteorological event are scarce (Cox *et al.* 2018).

67

68 This study compares responses of foraging seabirds to periodic tidal currents and an  
69 intermittent meteorological event in the Ria de Vigo, north-western Spain (42° 15' 04" N, 8°  
70 53' 30" W) (Fig.1). During the study, an estuarine plume (the Western Iberian Buoyant  
71 Plume, WIBP; Sousa *et al.* 2014) originating from the Minho Estuary (Fig.1) entered the ria,  
72 following an extreme river discharge event and southerly winds. In the same area, tidal  
73 currents flow through a narrow (2.8 km) and shallow (~25 m) channel (Fig.1), causing  
74 periodic variation in their direction/speed. This study uses the co-occurrence of these tidal  
75 currents and the WIBP intrusion to ask whether: (1) temporal changes in the number of  
76 foraging seabirds are correlated to these physical processes, (2) the strength of correlations  
77 are greater for physical processes associated with tidal currents or the WIBP intrusion?

78

## 79 **METHODS**

80

### 81 **Study Area**

82

83 This study was conducted on seven days between 4 and 15 June 2018. This period coincided  
84 with the breeding seasons of the dominant seabird species in the ria: European Shag  
85 *Phalacrocorax aristotellis* and Yellow-legged Gull *Larus michahellis*. The study area  
86 covered approximately 48 km<sup>2</sup> in the northern ria (Fig.1). The latter encompasses sand-banks  
87 known to be exploited by shags and gulls feeding predominantly on sandeel *Ammodytidae*  
88 (Velando *et al.* 1999) and Henslow's swimming crab *Polybius henslowii* (Munilla 1997),  
89 respectively. The recording of temporal variations in the numbers of foraging seabirds and  
90 physical processes occurred exclusively within the study area.

91

### 92 **Seabird Abundance**

93

94 A single observer recorded temporal variation in the number of foraging seabirds during 40  
95 zig-zag transects of approximately 10.3 km in length (Fig. 1). Transects were performed from  
96 a rigid inflatable boat moving at an average speed of 14 kt (11.2 – 17 kt), and lasted an  
97 average of 23 min (19 – 30 min). The numbers of transects were spread relatively evenly  
98 between ebb ( $n = 19$ ) and flood ( $n = 21$ ) tides. Throughout the transects, the observer  
99 followed European Seabirds At Sea (ESAS) methodology (Tasker *et al.* 1984). However, the  
100 observer was only 1 m above sea-surface. To ensure that the observer recorded representative  
101 numbers of animals, transects were only performed when the sea state was less than Beaufort

102 Scale 3. Nevertheless, estimations of sea state were recorded at the start of each transect to  
103 account for possible changes in the detectability of animals during rough weather  
104 (Camphuysen *et al.* 2004). These estimations represented a mean across the study area, and  
105 included non-integer values if there were spatial variations in weather conditions. Animals  
106 seen diving, dip-feeding and searching were considered as foraging seabirds (Camphuysen *et al.*  
107 *et al.* 2012). As transects were performed away from breeding colonies, animals sitting on the  
108 sea-surface were likely resting between foraging bouts rather than alongside nests (Waggitt *et al.*  
109 *et al.* 2016a, 2016b), and were also considered as foraging seabirds. Yellow-legged Gulls seen  
110 scavenging around fishing vessels were not considered as foraging seabirds (Valeiras 2003).

111

## 112 **Physical Processes**

113

114 Periodic tidal currents were quantified using outputs from an existing Finite Volume  
115 Community Ocean Model (FVCOM) (Chen *et al.* 2003) developed for the ria. Outputs were  
116 available at 15 min and approximately 100 m resolution. Mean depth averaged speed would  
117 summarise general conditions over the study area, whereas maximum or surface speeds may  
118 detect the presence or absence of strong hydrodynamic features at certain locations  
119 (Benjamins *et al.* 2015). Analyses was concerned with variations in the number of foraging  
120 seabirds across the study area, rather than associations between foraging seabirds and strong  
121 hydrodynamic features (e.g. Waggitt *et al.* 2016a). Therefore, for each transect, periodic tidal  
122 currents were represented by the mean depth averaged speed ( $\text{m s}^{-1}$ ) across the study area at  
123 the start of observations (Supplementary Material, S1). To discriminate between current  
124 directions, currents from the north were converted into negative values. Therefore, negative  
125 values show ebb currents, and positive values represent flood currents.

126

127 The WIBP intrusion was quantified using outputs from an existing Nucleus for European  
128 Modelling of the Ocean (NEMO) model (Madec 2008) developed for the Iberian region  
129 (Sotillo *et al.* 2015) (<http://marine.copernicus.eu>). Outputs were available at daily and 7 km  
130 resolution. For each transect, the influence of the WIBP was represented by the mean salinity  
131 (ppt) across the study area on the day of observations (Sousa *et al.* 2014). The arrival and  
132 departure of the WIBP intrusion in the study area would be identified by decreasing and  
133 increasing salinities, respectively. Data processing was performed in the ‘raster’ package  
134 (Hijmans 2013) in R (version 3.5.1, R Development Core Team 2018).

135

## 136 **Statistical Analysis**

137

138 Generalised Additive Models (GAMs) identified and quantified correlations between the  
139 number of foraging seabirds and physical processes (Wood 2006). A negative binomial  
140 distribution was used to account for overdispersion in the number of seabirds. The response  
141 variable was the number of foraging seabirds seen per transect. The explanatory variables  
142 were the corresponding measurements of depth averaged current speed, salinity and sea state.  
143 Salinity and sea state were modelled as continuous and linear variables. Whilst sea state is  
144 sometimes modelled as a categorical variable, a general decrease in detectability with  
145 increasing sea state was expected, making a linear variable more appropriate. Depth averaged  
146 current speed was modelled as a continuous and non-linear variable, with the number of  
147 knots fixed at 3. This setup allowed relationships with maximum speed, maximum speed in a  
148 particular direction (south or north), and slack water to be detected. Salinity was modelled as  
149 a continuous and linear variable. Sea state was included to account for possible decreases in  
150 the detectability of foraging seabirds in rough weather (Camphuysen *et al.* 2004). GAM were  
151 constructed using the ‘mgcv’ package (Wood 2006) in R.

152

153 Backwards model-selection based on *p*-values was performed (Zuur *et al.* 2009). Residuals  
154 from resultant models showed no evidence of temporal autocorrelation (Supplementary  
155 Material S2). Predicted variances in the number of foraging seabirds across gradients in  
156 physical processes were calculated from model parameters. In these calculations, the physical  
157 process of interest was varied between its minimum and maximum value, whilst other  
158 physical processes were held at their mean values. The magnitude and strength of  
159 relationships between numbers of foraging seabirds and physical processes were quantified  
160 using proportional differences (*Pd*). *Pd* represented the absolute difference between the  
161 maximum and minimum predicted values divided by the minimum predicted value, allowing  
162 direct comparisons between physical processes (Waggitt *et al.* 2017, 2018). Model selection  
163 and prediction were performed using the ‘mgcv’ package in R.

164

## 165 **RESULTS**

166

167 The WIBP intrusion originated from the Minho estuary following an extreme river discharge  
168 event on June 6. Southerly winds (see <https://www.meteogalicia.gal>) then advected the WIBP  
169 towards the study site between 7 and 9 June (Fig. 2). Decreasing salinities indicated the

170 arrival of the WIBP on June 10, with increasing salinities indicating its dispersal on June 12  
171 (Fig. 3). Periodic tidal currents were considerably faster when flowing from the north than  
172 from the south, with rapid changes in direction seen at slack water (Fig. 3).

173

174 The mean daily count of foraging European Shags peaked at 64.8 on 11 June, coinciding with  
175 the WIBP intrusion (Fig.3). The highest count in one transect on 11 June was 100 animals.  
176 On the remaining six days, the daily mean count was considerably lower. However, counts  
177 were generally higher before 11 June (lowest = 15.3, highest = 22.6) than after (lowest = 4.7,  
178 highest = 16.5) (Fig.3). The decrease after 11 June coincided with higher occurrence of rough  
179 weather; 71% of transects experienced sea states greater than Beaufort Scale 1.5.  
180 Accordingly, European Shags showed negative relationships with salinity and sea state  
181 (Fig.4). No relationships were found with depth averaged current speed. When accounting for  
182 the effect of sea state, *Pd* values indicated that (on average) 3.6 times more European Shags  
183 were encountered during WIBP intrusions than typical scenarios (Fig.4).

184

185 The daily mean count of foraging Yellow-legged Gulls also peaked on 11 June (56.3), again  
186 coinciding with the WIBP intrusion (Fig.3). The highest count in one transect on 11 June was  
187 94 animals. Daily mean counts after 11 June were comparable to those during the plume  
188 event (lowest = 34.3, highest = 44.6); those before were considerably lower (lowest = 9.0,  
189 highest = 16.5) (Fig.3). The former coincided with higher numbers of transects being  
190 performed in rough weather (see above) (Fig.3). Accordingly, Yellow-legged Gulls showed  
191 negative relationships with salinity, and positive relationships with sea state (Fig.4). No  
192 relationships were found with depth averaged current speed. When accounting for the effect  
193 of sea state, *Pd* values indicated that (on average) 4.2 times more Yellow-legged Gulls were  
194 encountered during WIBP intrusions than typical scenarios (Fig.4).

195

## 196 **DISCUSSION**

197

198 This study quantified the influence of periodic tidal currents and an intermittent  
199 meteorological event on the number of foraging seabirds in the Ria de Vigo in north-western  
200 Spain. Foraging European Shags and Yellow-legged Gulls showed no responses to periodic  
201 tidal currents. By contrast, numbers of both species increased during a WIBP intrusion on 11  
202 June. The numbers of foraging seabirds were also correlated to measurements of sea state.



203 The discussion focusses on responses of foraging seabirds to periodic tidal currents,  
204 intermittent meteorological events, and comparisons between these physical processes. The  
205 potential impacts from anthropogenic-driven changes within the ria are also discussed.

206

### 207 **Periodic Tidal Currents**

208

209 Increases in the numbers of foraging seabirds during certain tidal states are commonplace in  
210 areas of both strong ( $> 1 \text{ m s}^{-1}$ ) (Benjamins *et al.* 2015) and weak currents ( $< 0.5 \text{ m s}^{-1}$ )  
211 (Embling *et al.* 2012, Scott *et al.* 2013). The absence of responses to periodic tidal currents in  
212 the ria suggests that the amount of prey advected from surrounding areas is consistent across  
213 tidal states and/or turbulent eddies and shear-lines emerging during certain tidal states do not  
214 increase prey availability. Alternatively, limited numbers of surveys across different tidal  
215 states and/or strong responses of foraging seabirds to the WIBP intrusion could prevent  
216 responses to periodic tidal currents being detected. Extending studies over longer periods  
217 could investigate these possibilities further by increasing the number of surveys performed  
218 across different tidal states and outside WIBP intrusions. In any case, this study shows that  
219 strong tidal patterns in numbers of foraging seabirds cannot be assumed in coastal  
220 environments, even though they represent a conspicuous physical process.

221

### 222 **Intermittent Metrological Events**

223

224 Increased numbers of foraging seabirds in areas and seasons of persistent estuarine plumes  
225 are commonly reported (Cox *et al.* 2018). However, evidence of responses to an individual  
226 estuarine plume intrusion are scarce (Cox *et al.* 2018). As with previous examples, increases  
227 in the numbers of foraging seabirds during the WIBP intrusion are presumably explained by  
228 higher prey biomass. Local Yellow-legged Gulls forage primarily on Henslow's swimming  
229 crab (Munilla 1997). This detritivorous crab benefits from terrestrial-matter entering the  
230 water column (Vinagre *et al.* 2012), and observers noted Yellow-legged Gulls catching  
231 swarming crabs at the water surface. Whilst local European Shags forage consistently on  
232 sandeel, they sometimes exploit sand smelt *Atherina presbyter* in large numbers (Velando *et*  
233 *al.* 1999). This brackish-water fish (Wheeler 1969) is locally abundant, and it is speculated  
234 that European Shags exploited schools moving into the ria. However, whilst WIBP intrusions  
235 are commonplace in the ria (Des *et al.* 2019), studies over longer periods are needed to  
236 determine if responses occur during all WIBP intrusions.

237

238 Sea state is usually included in analyses to account for decreased detectability of animals  
239 during rough weather (Camphuysen *et al.* 2004). As expected, observers detected fewer  
240 European Shag in higher measurements of sea state. However, they detected more  
241 Yellowlegged Gull under the same circumstances. This could still indicate variation in  
242 detectability. The authors observed that Yellow-legged Gulls became restless during rough  
243 seas, and the tendency to take-off and land frequently could increase their detectability.  
244 However, it could also indicate differences in behaviour. European Shags detect and capture  
245 prey on the seabed using pursuit-dives. Therefore, European Shags may remain onshore  
246 during rough weather due to increased dive costs (Daunt *et al.* 2006, Lewis *et al.* 2015). By  
247 contrast, Yellow-legged Gulls detect and capture prey at the sea surface using dip-feeding or  
248 pecking. Animals could benefit from rough weather due to decreased flight costs (Haney *et*  
249 *al.* 1994) and resuspension of sub-surface material (Simpson *et al.* 2012). Therefore,  
250 relationships with sea state could be explained by both detectability and behaviour.

251

## 252 **Comparisons**

253

254 Periodic tidal currents are known to influence prey availability, initiating responses by  
255 foraging seabirds (Hunt *et al.* 1999, Benjamins *et al.* 2015) . However, this study shows that  
256 an intermittent meteorological event can cause stronger responses in some circumstances.  
257 These two processes almost certainly function synergistically, with numbers of foraging  
258 seabirds responding to the resultant conditions. Nevertheless, the relative influence of  
259 periodic tidal currents and intermittent meteorological events may relate to their control on  
260 conditions at a location. For instance, foraging Black-legged Kittiwakes *Rissa tridactyla*  
261 showed a greater response to periodic tidal currents in locations where current speeds were  
262 stronger (Trevail *et al.* 2019). Whilst the speed of periodic tidal currents cannot be considered  
263 weak in the ria, intermittent meteorological events have a much greater control on conditions  
264 in this area (Aristegui *et al.* 2006). This study suggests that the dynamics of foraging seabirds  
265 are intrinsically linked to that of the dominant process at a location.

266

## 267 **Anthropogenic Impacts**

268

269 The frequency of extreme river discharge events is likely to decrease in north-western Spain

270 (Cardosa Pereira *et al.* 2019). Studies over longer periods are needed to investigate responses  
271 to periodic tidal currents better, and whether responses to WIBP intrusions are commonplace.  
272 However, if WIBP intrusions consistently enhance prey availability, then observations of large  
273 numbers of foraging seabirds using the ria could become rarer. Moreover, if animals  
274 breeding/roosting in the ria depend on occasional WIBP intrusions for their subsistence, they  
275 could suffer from decreased prey encounters and increased searching efforts. This study shows  
276 that investigating responses to periodic tidal currents and intermittent meteorological events  
277 identifies potential impacts from anthropogenic-driven changes in coastal environments.

278

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280

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286

287

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289

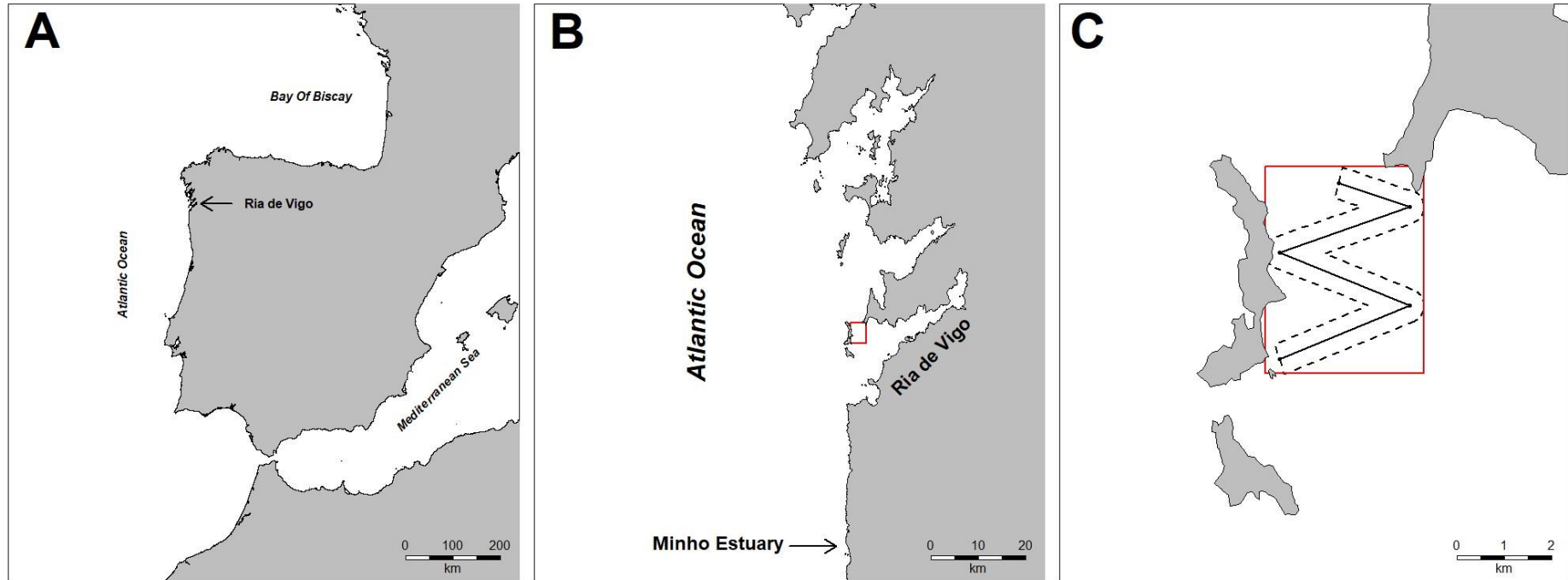
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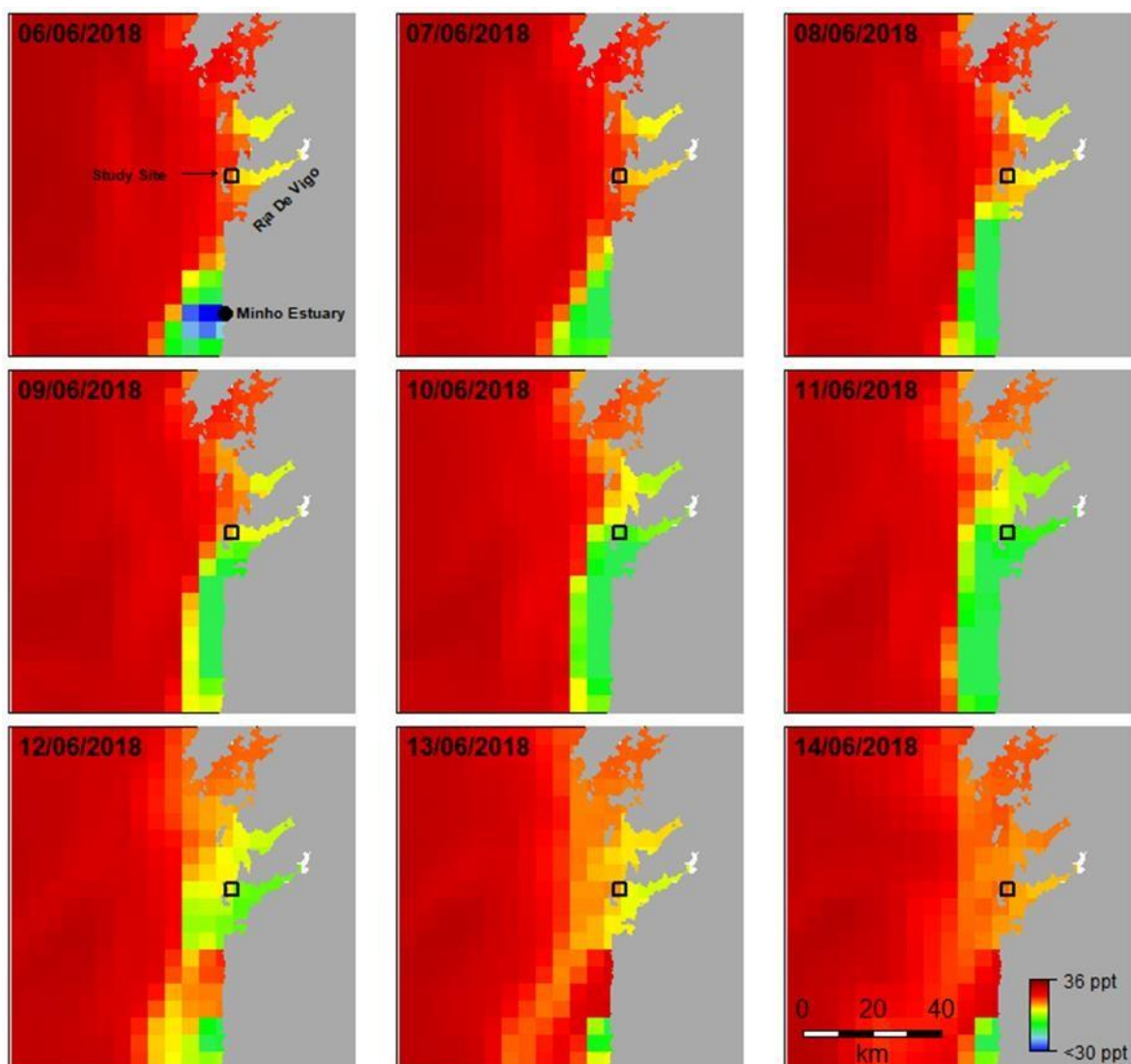
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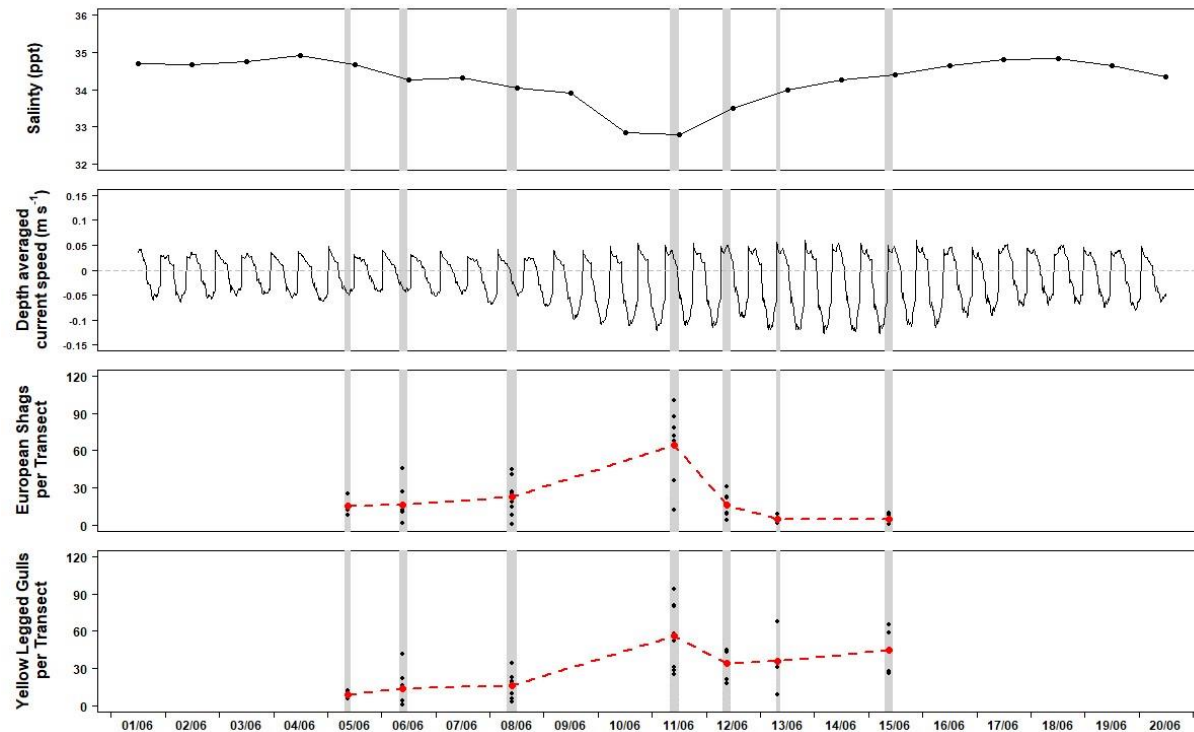


**Fig 1:** (A) The location of the Ria de Vigo in north-western Spain, (B) the area surrounding the ria, and (C) the zig-zag transects (solid black line) and observation area (dashed line) used to count numbers of foraging seabirds. The study area is shown by a red box.

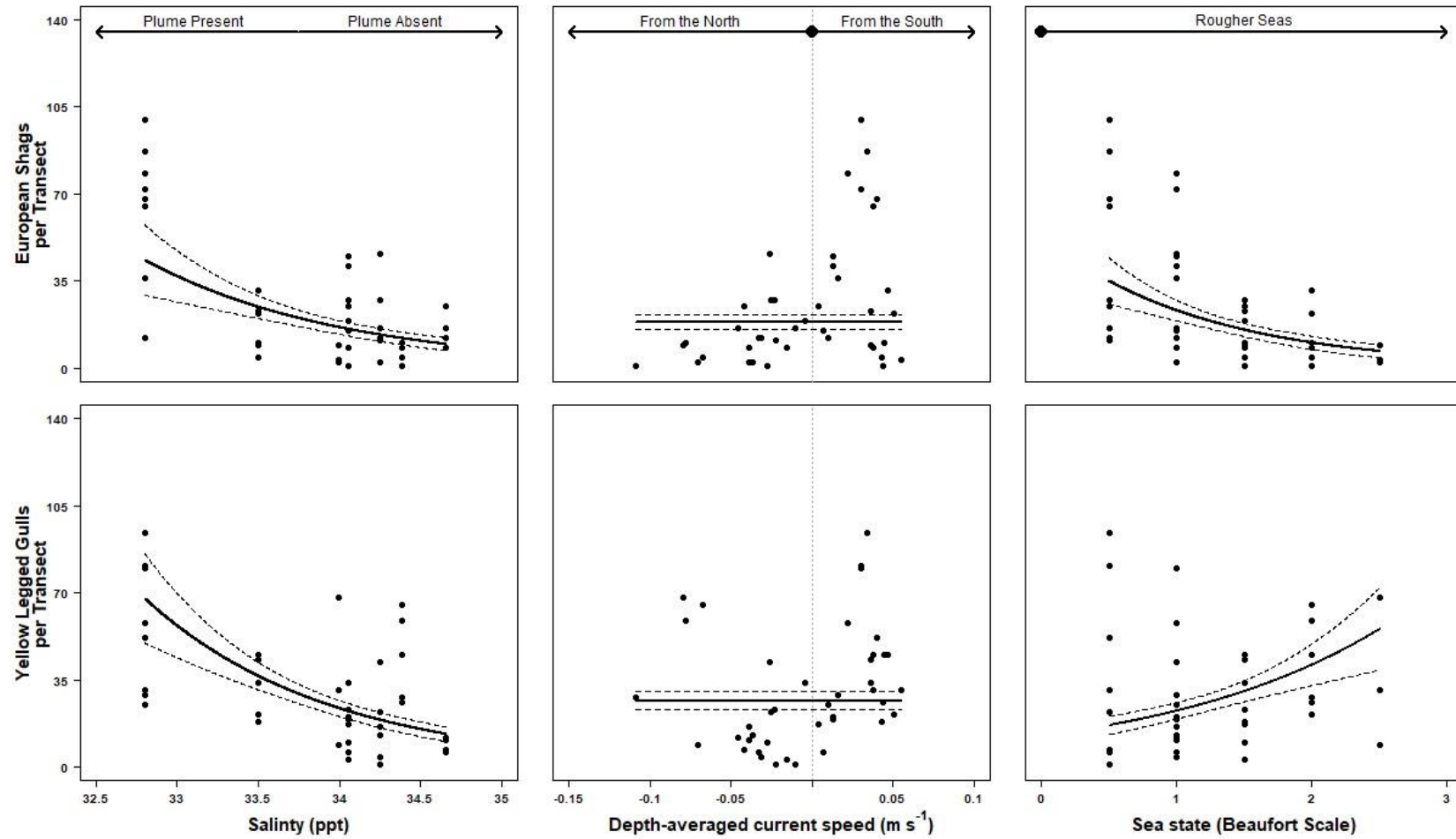




**Fig 2:** Variations in salinity between 6 and 14 June 2018 in the Ria de Vigo and the area surrounding the ria in north-western Spain. Values were sourced from an existing Nucleus for European Modelling of the Ocean (NEMO) model (Madec 2008) developed for the Iberian region (Sotillo *et al.* 2015). The study area is shown by a black box.



**Fig 3:** Temporal variations in salinity (ppt), depth-averaged tidal current speed ( $\text{m s}^{-1}$ ) and numbers of foraging seabirds during June 2018 in the Ria de Vigo, north-western Spain. Values of salinity were sourced from an existing Nucleus for European Modelling of the Ocean (NEMO) model (Madec 2008) developed for the Iberian region (Sotillo *et al.* 2015). Values of depth-averaged tidal current speeds were sourced from an existing FVCOM (Chen *et al.* 2003) developed for the ria. Negative values of depth-averaged tidal current speed represent flows from the north, whereas positive values represent flows from the south. Grey bars indicate times of zig-zag transects recording the numbers of foraging seabirds. Black points represent individual counts of foraging seabirds from zig-zag transects. Red points and lines illustrate daily mean counts of foraging seabirds among zig-zag transects.



**Fig 4:** Predicted variations ( $\pm$  standard error) in counts of foraging seabirds across different physical conditions from 4 to 15 June in the Ria de Vigo, north-western Spain. Predictions were made using generalized additive models (GAM) with a negative binomial distribution.