- 1 Environmental risk of trace elements in mangrove ecosystems: An assessment of
- 2 natural vs oil and urban inputs
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## Highlights

- City sewages and boatyards act as point sources for Zn, Cu and Pb.
- Ni and V contents show that oil industry has a low impact in mangroves.
- Zn, Cu, and Pb contents exceed adverse effect index (AEI) in two urban point sources.

## Abstract

- The petrochemical industry and urban activities are widely recognized worldwide as a source of pollution to mangrove environments. They can supply pollutants such as trace elements that can modify the ecosystem structure and associated services, as well as human populations. Through geochemical data, multivariate statistical analysis and pollution indices such as the enrichment factor (EF), geo-accumulation index (Igeo), adverse effect index (AEI) and the pollution load index (PLI), we evaluated the factors that control trace element distribution, punctual sources and determined the pollution level of sediments and their potential biological impact in the mangrove ecosystem of Isla del Carmen, Mexico. The factor and cluster analysis highlighted that the distribution of trace elements is influenced by the mineralogy, texture as well as urban derived sources. The pollution indices showed values in the punctual sources from the urban area of EF > 10, Igeo > 3, AEI >3, PLI > 1 by Cu, Zn and Pb. Finally, the results revealed that mangroves from Isla del Carmen has a major influence from urban activities and natural sources rather than oil industry and also indicate a degraded environment as a result of anthropogenic activities that could have knock-on effect for human health if polluted marine organisms derived from the urban mangroves are consumed.
- Capsule abstract: Surface sediments show the influence of point sources on selected trace
  element concentrations correlated with human activities within the mangroves of Isla del
  Carmen, Mexico.
- **Keywords:** Trace Elements, Sediment, Oil Industry, Mangroves, Gulf of Mexico, Isla del Carmen

#### 1. Introduction

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Mangroves are vegetated intertidal coastal environments located in tropical and subtropical areas. They are characterized by the presence of plants adapted to frequent inundation by saline water imposed by a tidal regime (NOAA, 2014). To support and sustain the well-being of human being, mangroves provide a range of ecosystem services such as: nursery and breeding habitats for commercial shrimp, crabs and fish species; coastal protection against erosion processes, cyclones, storms, floods and carbon sequestration (Brander et al., 2012; Lovelock et al., 2015; Mehaur et al., 2018; Shi et al., 2019). Like many other ecosystems, mangroves have been subjected to anthropogenic pressures resulting in decreases in mangrove area, particularly conversion to agriculture, aquaculture and salt ponds, urban expansion as well as alterations to sediment supply, sea level rise, drought and storms as a result of climate change (Ward et al., 2016; Veetil et al., 2018; Mafi-Gholami et al., 2019). Due to their location in coastal and estuarine settings they are also often exposed to a range of pollutants particularly those located proximal to industry or urban settings. Many of these pollutants are harmful to mangrove associate organisms and humans when transferred through the trophic chain (Senthilkumar et al., 2013; Kulkarni et al., 2018; Chai et al., 2019). Trace elements are common pollutants found in mangroves and are considered as a serious threat to human health and other living organisms due to their persistence, potential for bioaccumulation, and toxicity (Arrivabene et al., 2016; Celis et al., 2017; Singh and Kumar, 2017). However, it is important to note that trace elements can be derived from anthropogenic and natural sources that complicate analysis concerning their environmental impact (Anaya-Gregorio et al., 2018; Armstrong-Altrin et al., 2019). To resolve this issue geochemical indices such as the enrichment factor (EF), geoaccumulation Index (Igeo), pollution load index (PLI), adverse effect index (AEI), as well as statistical tools have been developed to provide a better understanding about the environmental risks that these

- substances pose to mangrove ecosystems and humans (Goher et al., 2014; Duodu et al., 2017;
- 73 Benson et al., 2018).
- 74 Mangroves from Isla del Carmen are likely to be influenced by point source pollution from
- 75 urban, industrial and petrochemical derived sources as a result of the proximity to an industrial
- 76 port city and petrochemical source. Therefore, Isla del Carmen is an important area to study
- how oil industry and urban activities impact mangrove environments.
- 78 The aim of this study is to determine the agents that influence trace element distribution in
- 79 mangroves with pollution sources including the oil industry and urban activities. Additionally,
- 80 we aimed to evaluate the contamination level of sediments and their potential biological
- 81 impact in the mangroves of Isla del Carmen, Mexico.

# 2. Study area

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Isla del Carmen is located in the southeastern Gulf of Mexico in Campeche, Mexico (Figure 1). Twenty three percent of its surface corresponds to the urban area of Ciudad del Carmen, while the remaining 77% is covered by mangroves (INEGI, 2018). Ciudad del Carmen is the second most populated urban center (248,303 inhabitants) in Campeche state (INEGI, 2018). Isla del Carmen belongs to a natural protected area known as Wildlife Protection Area Laguna de Terminos, which includes the largest coastal lagoon environment in Mexico (about 7050.16 Km²) (INEGI, 2018). In the north coast of the island the marine environment dominates and houses the Canterell oil field, which is the largest gas (38%) and offshore oil (56.5%) producer in Mexico (Nava et al., 2018; PEMEX, 2018) while, in the south coast it receives fluvial discharges of the Palizada, Chumpan and Candelaria rivers, which supply nutrients, sediments, as well as pollutants generated by agricultural, cattle feeding and other human activities (Carvalho et al., 2009). The weather is mostly humid warm with summer rainfall. The average annual temperature ranges from 26 to 28 °C, while the average annual

rainfall ranges from 1500 to 2000 mm (INEGI, 2018). Seasonally, three different regimes are recognizable, the dry season (Feb-May), rainy season (Jun-Sep), and stormy season (Oct-Jan).

## 3. Materials and methods

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Sediment samples were collected from thirty-six mangrove sites on Isla de Carmen, Mexico in August 2019 that corresponded to the rainy season; this season was chosen because the rivers' influence is higher in the area rather than other seasons (Figure 1). Samples were collected from creek edges directly adjacent to the mangroves using a Van Veen dredge from a boat at a water column depth from 0.5 to 2.5 m at mid-high tide to avoid running aground. In order to avoid contamination during sampling sediment samples were removed with a plastic spatula from the middle part of the dredge in such a way that sediments had no contact with the metallic parts of the equipment and to get the most recently deposited sediment, we collected the samples in first centimetre. Each sample was stored in sample bags at 4°C until chemical analysis. Samples were divided in two halves with one half used for geochemical analysis and the other for granulometry. Samples were dried at 50 °C for 5 days and homogenized in an agate mortar by hand. All concentrations in this work were expressed with reference to dry weight. Major and trace element concentrations were determined using a RIGAKU ZSX Primus II X-ray fluorescence spectrometer system and analyzed in pressed powder briguettes. The accuracy of major elements was evaluated using the rock standard Argillite whose values ranged from 102.6 to 93%, with the exception of  $Na_2O$  (70.0%),  $P_2O_5$  (73.1%) and  $K_2O$  (89.3%). The accuracy of trace elements was estimated using the standard CH-1 marine sediment, values ranged from 102.6 to 90.8% with the exception of V (111.8%) and Co (113.8%). Sediment particle size distribution was determined by sieving with a RX-29 Ro-Tap Sieve Shaker and a standard ASTM sieve set from -2.0, -1.0, 0.0, 1.0, 2.0, 3.0, and 4.0 Φ. Sediments were classified as mud, sand, and gravel

based on the methodology proposed by Folk (1980). Organic matter and carbonates were

estimated using a loss on ignition method: 1 g of sediment samples were weighed before and after incineration in a muffle furnace at 550 °C and again at 950 °C for 24 hrs in order to eliminate the organic matter and carbonates (Ahmed et al., 2018). Total organic matter and carbonates are reported as percentage of dry weight.

The results were interpreted using descriptive and multivariate statistical tools including correlation analysis to evaluate the strength of relationship among geochemical parameters. A cluster analysis was used to identify impacted locations in mangroves from Isla del Carmen, and factor analysis was performed to determine the variables that better explain the system variability (Statistical analyses were performed using the software STATISTIC 8). Geochemical indices such as EF, Igeo, PLI, and AEI were employed to evaluate the natural or anthropogenic origin for each trace element: to compare and contrast environmental quality with other locations inside the island, and to infer any possible adverse effect on benthic biota (Celis et al., 2018).

**Enrichment Factor** 

135 This proxy is evaluated using equation (1)

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$$EF = \frac{\left(\frac{X_{sample}}{Y_{sample}}\right)}{\left(\frac{X_{background}}{Y_{background}}\right)} \dots \dots \dots \dots (1)$$

Where  $X_{sample}$  and  $X_{background}$  are the concentrations of any trace element analysed in the sample and  $Y_{sample}$  and  $Y_{background}$  are the concentrations of a conservative element that is used to normalize the data. In this work, we used  $Al_2O_3$  as a conservative element because it represents the clay fraction in the sediment and it is not affected by anthropogenic sources or geochemical changes such as redox processes (Celis et al., 2013). As background we employed the least contaminated station (Feng et al., 2010) rather than the Upper Continental Crust (UCC) (McLennan, 2001) to avoid any overestimation in the data interpretation. The

- enrichment factor results were divided into ranks to establish the degree of enrichment (Birth,
- 145 2003) (Table S2).
- 146 Geo-accumulation index
- 147 This proxy is evaluated using equation 2 (Müller, 1969).

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$$I_{geo} = log_2 \left( \frac{C_n}{1.5B_n} \right)$$
 (2)

- 149 Where "C<sub>n</sub>" represents the concentration of any trace element in the sediment sample and
- 150 "B<sub>n</sub>" represents the background concentration of the same element; the value of 1.5 is a factor
- that considers possible variability generated by lithological changes. Geo-accumulation index
- results were divided into ranks to establish the degree of pollution (Birth, 2003) (Table S2).
- 153 Adverse Effect Index

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- The adverse effect index (AEI) can be applied to evaluate how many times the metal concentration in the sediment has exceeded the Threshold Effect Level (TEL) developed by Long et al., (1995); as well as inferring whether trace element concentrations in the sediment could adverse impact benthic biota (Muños et al., 2012; Hamdoun et al., 2015; Baptista et al., 2017). An AEI lower than 1 means that the trace element concentration in the samples is not high enough to produce adverse effects from sediment contamination to organisms where no
- 161 This proxy is calculated using equation (3)

toxicity studies are undertaken or available.

$$AEI = \frac{[MC]}{[SQG_s]}$$

Where MC is the trace element concentration in the sample and SQGs is any sediment quality guideline such as Effect Range Low (ERL), Threshold Effect Level (TEL) and others. SQGs are based on the collection, revision and the integration of a range of studies performed in North

America using laboratory bioassays, equilibrium-partitioning modelling and field studies on the toxicity of metals in sediment on benthic composition (Long et al., 1995). In this work, we used the Threshold Effect Level (TEL) and the Analogous Effect Threshold (AET). Both were obtained via the Screening Quick Reference Tables (SQuirRts) supplied by NOAA, 2016.

Pollution Load Index

The pollution load index is calculated using equations 4 and 5 (Tomlinson et al., 1980).

$$PLI_{sample} = \sqrt[n]{CF_{metal 1} * CF_{metal 2} * ... CF_{metal n}} ... ... ... ... ... (4)$$

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$$CF = \left[\frac{C_{mS}}{C_{mB}}\right]$$
.....(5)

In general; the PLI is calculated with the  $n^{th}$  root of the variable called contamination factor (CF) which is evaluated using the equation (4). The super index "n" in the root represents the total number of samples. While the CF is defined as the ratio between the element concentration analyzed in the sample " $C_{ms}$ " and the background concentration of the same element that is identified in the equation (5) as " $C_{mB}$ ". The background was derived using the least contaminated station (Feng et al., 2010). According to Tomlinson et al., (1980), PLI values of zero suggest the absence of pollutants, while PLI values of one or greater denotes the presence of pollutants or progressive deterioration of sediment quality.

#### 4. Results

# 4.1. Textural analysis

Mud (silt and clay) content in the samples varied from 26.5 to 92.1%; while sand and gravel varied from 6.0 to 67.0% and 0.05 to 40.3%, respectively (Table 1). The mud, sand and gravel ternary diagram (Figure 2) showed that 75.1% of the samples were classified as mud, 13.8% were classified as muddy-sand and 11.1% were not well classified because they had similar values for each of these parameters.

#### 4.2. Geochemical composition

The average organic matter (OM) content in the study area was 18.1% and varied between 6.8 and 36.4%. The highest concentration was found at station 6, which was located within the urban area of Ciudad del Carmen (Figure 1). The lowest concentration was found at station 31, approximately 35 km from station 6, such a contrast suggests that organic carbon could be associated to the local drainage piping of Ciudad del Carmen. The average carbonate concentration was 21.2%. The highest carbonate values were found in the middle of the island at stations 25 (32.4%), 26 (37.5%) and 27 (33.9%), which are characterized by the presence of shell banks and higher energy conditions. The lowest values where found at station 11 (8.3%) and 21 (6.5%). Major element concentrations varied within the following ranges: 11.4 – 35.3% for CaO, 11.2 – 34.5% for SiO<sub>2</sub>, 2.23 – 9.53% for Al<sub>2</sub>O<sub>3</sub> and 1.49 – 4.52% for Fe<sub>2</sub>O<sub>3</sub> (Table 1). The range and average of trace elements are shown in Table 1. Stations 7 and 1 that are located close to the city sewages exhibited the highest values for Zn (794 mg kg<sup>-1</sup>, 365 mg kg<sup>-1</sup>), Cu (311 mg kg<sup>-1</sup>, 81mg kg<sup>-1</sup>) and Pb (111 mg kg<sup>-1</sup>, 48 mg kg<sup>-1</sup>) (Figure 3, Figure S2). Station 10 situated close to a boat yard, where there is limited water flux because of mangroves exhibited high values of Zn (471 mg kg<sup>-1</sup>), Cu (35 mg kg<sup>-1</sup>) and Pb (34 mg kg<sup>-1</sup>) (Figure 3, Figure S2).

#### 4.3. Pollution indices

# 4.3.1. Enrichment Factor

All the trace elements showed variations in their enrichment level degrees (Figure 4). All samples exhibited levels from no enrichment to minor enrichment for V, Cr, Rb, Sr, Ba and Zr. Moderate to severe EFs were found for Co and Ni at station 30 (5.1) and 30 (5.3) respectively. Values of EFs > 10 (severe and very severe enrichment) were found for Cu at stations 1 (24.5) and 7 (63.1), for Zn at stations 1 (27.6), 7 (40.3) and 10 (12.2), and for Pb at stations 1 (14.5) and 7 (22.5).

#### 4.3.2. Geo-accumulation Index

The geo-accumulation Index shows (Figure 4) that none of the samples were moderately polluted with Sr, Ba and V. Between 4 and 14 of the samples were moderately polluted with Cr, Zr, Zn, Ni, Rb, Cu, Pb and Co (Figure 4). The highest values of Igeo were found at stations 1 (4.2, 6.3, 3.4), 7 (6.1, 7.5, 4.6), 10 (3, 6.7, 2.9), 29 (3.4, 4.7, 1.7) and 30 (1.6, 3.7, 1.2), being moderately to extremely polluted with Cu, Zn and Pb, respectively.

#### 4.3.3. Adverse Effects Index

The estimated AEI values for the mangroves of Isla del Carmen showed that the samples could be associated with adverse effects on organisms caused by Cr (100%), Ba (100%), Ni (47.2%), Co (25%), Cu (11.1%), V (8.3%) and Zn (8.3%), (Figure 4). The *maximum* AEI values were found at station 7 for Cu (16.6), Zn (6.4), Pb (3.7) and Ba (3.2). While the *maximum* values of AEI at station 11 corresponded to V (1.2), and Ni (5.1), at station 9 the highest AEI corresponds to Co (1.9) and they are the at stations located within a small boat channel next to the urban mangroves.

#### 4.3.4. Pollution Load Index

Calculated PLI values of > 1 were recorded at every site (Figure 5), which means that 100% of the samples had a significant deterioration of sediment quality due to trace element concentrations. However, for this index, there was an evident tendency of samples collected inside the tidal channels, and around Ciudad del Carmen, to exhibit PLI values > 2 (except station 13). Meanwhile, samples collected at places far away from human activity showed PLI values < 2 (except stations 29 and 30).

#### 4.4. Factor and cluster analysis of sediment samples

A factor analysis was performed to identify the causes of variation between the geochemical and textural data from the mangroves. Four factors were identified that explained 81% of the system's *variance*. All factors were associated with the mineralogy, texture, anthropogenic and

natural sources (Table 2). The first factor explained 29% of the total variance and was related with sediment mineralogy due to its interactions between major elements, carbonates and gravel. The second factor explained 20% of the total variance, and it was related to sediment texture because this factor grouped parameters such as sand, mud, OM, CaO, Na<sub>2</sub>O, and S, which are widely associated to a specific sediment grain size. The third factor explained 17% of the total variance and they grouped P<sub>2</sub>O<sub>5</sub>, Cu, Zn, and Pb. It was linked to anthropogenic sources because the pollution indices noted that these elements are supplied by anthropogenic and point sources. Finally, the fourth factor explained 15% of the total variance; it grouped V, Co, Ni and Rb, which was linked to natural sources because according to the pollution indices these elements were likely to be supplied by the Palizada, Chumpan and Candelaria river catchments lithology.

In order to identify associations between 36 sediment samples, a cluster analysis was performed. The cluster diagram showed three main groups (A, B and C) (Figure 6). Group A was characterized by a higher content of gravel, sand, OM and CO<sub>3</sub>. Group B was characterised by higher concentrations of SiO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>, K<sub>2</sub>O, MgO, V, Cr, Ni, Rb, Zr. Group C was characterised by high mud content, OM, Fe<sub>2</sub>O<sub>3</sub>, P<sub>2</sub>O<sub>3</sub>, S, Co, Cu, Zn, Pb, and Ba. Group C also showed the highest PLI values and highest EF, Igeo and AEI of Co, Cu, Zn, Pb and Ba.

## 5. Discussion

## 5.1. Major and trace elements source

The geochemical composition and the ternary diagram (Figure S1) showed that mangrove sediments in the area are richer in CaO (Carbonates) than  $Al_2O_3$  (aluminosilicates) and  $Fe_2O_3$  (heavy minerals). Carbonates are supplied by shell fragments of organisms that live in the mangrove such as snails, oysters and other mollusks and bivalves. The negative *correlation* between CaO and CO<sub>3</sub> (0.48, p < 0.05) suggests that although the study area is characterized by biogenic sediment rich in carbonates, there are other calcium sources such as plagioclases,

otherwise this *correlation* would be high. The negative *correlation* between CaO and MgO (-0.74, p < 0.05) noted that Mg substitutes Ca from carbonate minerals due to its similar radio atomic size. The positive *correlation* between Fe<sub>2</sub>O<sub>3</sub> and TiO<sub>2</sub> (0.95, p < 0.05) is related to heavy minerals. Positive *correlation* between Al<sub>2</sub>O<sub>3</sub> with SiO<sub>2</sub> (0.90, p < 0.05), and K<sub>2</sub>O (0.93, p < 0.05), highlighted the presence of aluminosilicates such as plagioclases and the positive *correlation* between Al<sub>2</sub>O<sub>3</sub> with SiO<sub>2</sub> (0.90, p < 0.05), Fe<sub>2</sub>O<sub>3</sub> (0.92, p < 0.05), TiO<sub>2</sub> (0.89, p < 0.05), and K<sub>2</sub>O (0.93, p < 0.05) suggests that all of these minerals are in fine particles such as mud rather than sand and gravel. On the other hand, negative *correlations* were recorded between CO<sub>3</sub> and SiO<sub>2</sub> (-0.88, p < 0.05), Al<sub>2</sub>O<sub>3</sub> (-0.82, p < 0.05), Fe<sub>2</sub>O<sub>3</sub> (-0.75, p < 0.05), TiO<sub>2</sub> (-0.75, p < 0.05), K<sub>2</sub>O (-0.88, p < 0.05), indicating that biogenic sourced sediments were diluted by terrigenous sediment from the Palizada, Chumpan and Candelaria rivers and it is supported by the ternary diagram (Figure S1) where it is clear how CaO concentration decreases when sediment is getting rich in Al<sub>2</sub>O<sub>3</sub>.

The trace element data and the cluster analysis (Figure 6) highlighted stations 7, 10 and 1 as punctual sources, in these stations the highest trace elements concentration and pollution indices of Zn, Cu and Pb are identified. The lack of positive *correlation* for these elements with almost all the studied parameters and their spatial distribution support the idea that the city sewages and boat yards act as anthropogenic sources for these elements (Table S1). In contrast, the positive *correlations* for V (0.59, 0.61, 0.57, 0.87, p<0.05), Ni (0.65, 0.60, 0.56, 0.86, p<0.05), with Al<sub>2</sub>O<sub>3</sub>, Fe<sub>2</sub>O<sub>3</sub>, TiO<sub>2</sub> and Co suggest a terrestrial natural origin probably supplied by Palizada, Chumpan and Candelaria rivers, where their lithology are linked to mafic rocks (Ortiz et al., 2006). Relatively low *correlations* between Ba and Zr with almost all major and trace elements (Table S1) suggest an oil industry origin; as these elements are likely to be related to crude oil, residual fuel oils, accidental oil spills and oil drilling (Fieldler et al., 2009; Zhang et al., 2015). On the other hand, the cluster analysis showed that the association between stations were divided by three geochemical groupings according to their mineralogy,

texture and anthropogenic/natural influence. This is supported by the factor analysis results where the mineralogy, texture, anthropogenic and natural sources were identified as factors that explained the geochemical variations from the mangrove sediments of Isla del Carmen.

## 5.2. Evidence of oil influence in the Carmen Island Mangroves

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V and Ni are elements that are related to oil industry because they are abundant in crude oil. Generally, the ratio Ni/V is used to identify oil families such as Brent, West Texas, Maya, etc (Barwaise, 1990). However, this ratio also has been used as an oil pollution index in places with presence of an important oil industry such as the Gulf of Mexico (Cuevas et al., 2018; Ruiz et al., 2019). We used this ratio to identify the oil industry influence on the mangroves of Isla del Carmen. According to Barwaise (1990) the content of Ni and V in crude oil is about 340 mg kg<sup>-1</sup> and 1580 mg kg<sup>-1</sup> respectively. Also, according to the IIE (1998), the chemical composition of the Mexican fuel oil could have up to 20-60 and 200-350 mg L<sup>-1</sup> of Ni and V respectively, giving two interval ratios from 0.3 to 0.17 and 0.1 to 0.05 evidencing geochemical markers that Ni and V come from Mexican oil industry activities. The ratios found in the mangroves of Isla del Carmen ranged from 0.5 to 1.5. Such values are higher than the values reported average in Veracruz (0.28) and Tamaulipas (0.17 - 0.23), states with high oil industry activity (Ruiz et al., 2012; Celis et al., 2018). These results suggest a low impact from oil industry sources and confirm the predominantly natural source of these trace elements. Recent studies on coastal sediments near to the Isla del Carmen also reported that the Ni and V contents are due to the natural sources and were derived by the contribution of intermediate and mafic source rocks like andesite and basalt (Armstrong-Altrin et al., 2018; Ramos-Vázquez and Armstrong-Altrin, 2019).

## 5.3. Ecological status of the area

The pollution indices such as the PLI reported in this study suggest that the Isla del Carmen mangroves are considered a degraded environment as a result of the trace elements present

in the sediments. The EF, Igeo and AEI highlighted specific sites close to point sources that caused severe enrichment, very severe enrichment and those sites were classified as strongly polluted and extremely polluted mangrove environments for Zn, Cu and Pb.

The adverse effects that are likely to be evident in the mangrove ecosystem as a result of these trace element concentrations would be bioaccumulation and biomagnification in species at all the trophic levels (Baki et al., 2018; Karar et al., 2019). In addition, because of their toxicity, mangrove organisms would show a reduction in their growth development, as well as sexual and genetic anomalies (Souza et al., 2015), and even there would be a decrease in the number of benthic species (crabs, clams), fish and birds (Souza et al., 2018; Garriz et al., 2019) that coexist in the mangrove environment. Also, it could affect other proximal coastal environments such as the Terminos Lagoon, and the estuarine environment that exists in this natural system. Finally, it could affect human health if polluted marine organisms derived from the mangroves are consumed by the human population; as these levels of the trace elements are likely to cause serious damage to the kidney, central nervous system, liver, intestinal tract and reproductive system (Siddique et al., 2012; Abdel-Khalek et al., 2016). However, it is necessary to conduct further investigation to assess the impacts of these pollutants on local ecological diversity and biodiversity to fully assess the ecological status in the area.

## 6. Conclusion

The bulk geochemical data noted that almost all the trace element concentrations in the mangroves from Isla del Carmen are close to background levels; except for Zn, Cu and Pb where the spatial distribution suggests that the city sewages and boatyards act as point sources for these elements. The trace element distributions in the mangroves from Carmen Island are driven by natural and anthropogenic factors such as the mineralogy, texture and urban activities or urban sources. Finally, the mangroves from the Carmen Island can be considered impacted by the wastes generated within the urban area rather than by the

extraction of oil activities that take place close to the Carmen Island, which suggests that there is limited impact from oil activities in the local area. However, levels are such that these areas can be considered as heavily polluted by some of the trace elements investigated and this is likely to cause adverse effects on benthic organisms associated with the mangroves.

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## References

- 350 Abdel-Khalek, A.A.A., Elhaddad E., Mamdouh S., Assm, M.S.M., 2016. Assessment of Metal
- 351 Pollution around Sabal Drainage in River Nile and its Impacts on Bioaccumulation Level, Metals
- 352 Correlation and Human Risk Hazard using Oreochromis niloticus as a Bioindicator. Turk. J. Fish.
- 353 Aquat. Sci. 16, 227-239.
- 354 Ahmed, I., Mostefa, B., Bernard, A., Oliver, R., 2018. Levels and ecological risk assessment of
- 355 heavy metals in surface sediments of fishing grounds along Algerian coast. Mar. Pollut. Bull.
- 356 136, 322-333.
- 357 Anaya-Gregorio, A., Armstrong-Altrin, J.S., Machain-Castillo, M.L., Montiel-García, P.C., Ramos-
- 358 Vázquez, M.A., 2018. Textural and geochemical characteristics of late Pleistocene to Holocene
- 359 fine-grained deep-sea sediment cores (GM6 and GM7), recovered from southwestern Gulf of
- 360 Mexico. J. Palaeogeog. 7(3), 253-271.

- 361 Armstrong-Altrin, J.S., Ramos-Vázquez, M.A., Zavala-León, A.C., Montiel-García, P.C., 2018.
- 362 Provenance discrimination between Atasta and Alvarado beach sands, western Gulf of Mexico,
- 363 Mexico: Constraints from detrital zircon chemistry and U-Pb geochronology. Geol. J. 53(6),
- 364 2824-2848.
- 365 Armstrong-Altrin, J.S., Botello, A.V., Villanueva, S.F., Soto, L.A., 2019. Geochemistry of surface
- 366 sediments from the northwestern Gulf of Mexico: implications for provenance and heavy
- metal contamination. Geol. Quarter. 63 (3), 522-538.
- 368 Arrivabene, H., Quenupe, C., da Costa, I., Wundorlin, D., Rozindo, D., Rodrigues, S., 2016.
- 369 Differential bioaccumulation and translocation patterns in three mangrove plants
- experimentally exposed to iron: Consequences for environmental sensing. Environ. Pollut. 215,
- 371 302-313.
- Baki, M.A., Hossain, Md., Akter, J., Quraishi, Sh. B., Shohib, Md. F.H., Ullah, A.K.M.A., Khan, Md.
- 373 F., 2018. Concentration of heavy metals in seafood (fishes, shrimp, lobster and crabs) and
- human health assessment in Saint Martin Island, Bangladesh. Ecotoxicol. Environ. Saf. 159,
- 375 153-163.
- Baptista, J.A., Ferreira, C., Gutterres, C., Monteiro, E., Vaz, G., Monica, O., 2017. Environmental
- 377 change in Guanabara Bay, SE Brazil, based in microfaunal, pollen and geochemical proxies in
- 378 sedimentary cores. Ocean Coast. Manag. 143, 4-15.
- 379 Barwise A.J.G., 1990. Role of Nickel and Vanadium in petroleum Classification. Energy & Fuels.
- 380 4, 647-652.
- 381 Benson, N.U., Adedapo, A.E., Fred-Ahmadu, O.H., Williams, A.B., Udosen, E.D., Ayeyuyo, O.O.,
- Olajire, A.A., 2018. New ecological risk indices for evaluating heavy metals contamination in
- aquatic sediment: A case study of the Gulf of Guinea. Reg. Stud. Mar. Sci. 18, 44-56.

- Birth, G., Woodcoffe, C.D., Furness, R.A., 2003. a scheme for assessing human impacts on
- 385 coastal environments using sediments. Coastal GIS 2003. Wollongong University Papers in
- 386 Centre for Marine Policy. 14 (Australia).
- 387 Brander, L.M., Wagtendonk, A.J., Hussain, S.S., McVittie, A., Verburg, P.H., de Groot, R.S.,
- Vander Ploeg, S., 2012. Ecosystem service values for mangroves in Southeast Asia: a meta-
- analysis and value transfer application. Ecosyst. Serv. 1, 62-69,
- 390 Carvalho, F.P., Villeneuve, J.P., Cattini, C., Rendon, J., Mota de Oliveira J., 2009. Pesticide and
- 391 PCB residues in the aquatic ecosystems of Laguna de Terminos, a protected area of the coast
- of Campeche, Mexico. Chemosphere 74, 988-995.
- 393 Celis, O., Rosales, L., Carranza, A., 2013. Heavy metal enrichment in surface sediments from
- the SW Gulf of Mexico. Environ. Monit. Assess. 185, 8891-8907.
- 395 Celis, O., Rosales, L., Cundy, A., Carranza, A., 2017. Sedimentary heavy metal(loid)
- 396 contamination in the Veracruz shelf, Gulf of Mexico: A baseline survey from a rapidly
- developing tropical coast. Mar. Pollut. Bull. 119, 204-213.
- 398 Celis, O., Rosales, L., Cundy, A., Carranza, A., Croudace, I., Hernandez, H., 2018. Historical trace
- 399 element accumulation in marine sediments from the Tamaulipas shelf, Gulf of Mexico: An
- assessment of natural vs anthropogenic inputs. Sci. Total Environ. 622-623, 325-336.
- 401 Chai, M., Li, R., Ding, H., Zan Q., 2019. Occurrence and contamination of heavy metals in urban
- 402 mangroves: A case study in Shenzhen, China. Chemosphere 219, 165-173.
- 403 Cuevas, H., Rosales, L., Marquez, A.Z., Carranza, A., 2018. Environmental assessment of a
- 404 mangrove protected area in the southeast Gulf of Mexico through sediment core analysis.
- 405 Environ. Earth Sci. 77, 73.

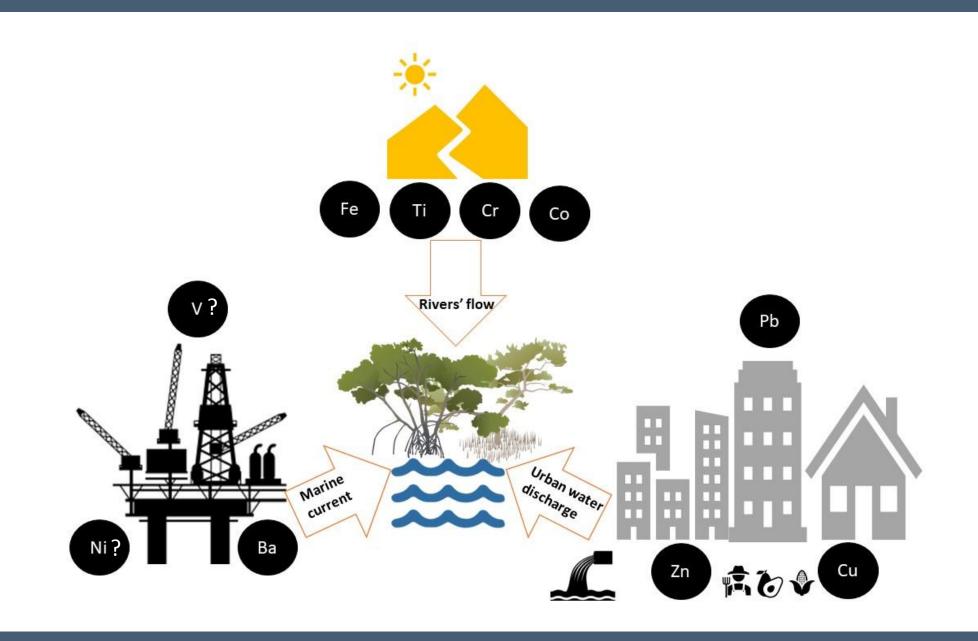
- 406 Duodu, G.O., Goonetilleke, A., Ayoko, G., 2017. Potential bioavailability assessment, source
- 407 apportionment and ecological risk of heavy metals in the sediment of Brisbane River Estuary,
- 408 Australia. Mar. Pollut. Bull. 117, 523-531.
- 409 Feng, H., Jiang, H., Gao, W., Weinstern, M., Zhang, Q., Zhang, W., 2010. Metal contamination in
- sediments of western Bohay Bay and adjacent estuaries, China. J. Environ. Manage. 34, 1-13.
- 411 Fiedler, S., Siebe, C., Herre, A., Roth, B., Cram, S., Stahr, K., 2009. Contribution of oil industry
- 412 activities to environmental loads of heavy metals in Tabasco low lands, Mexico. Water Air Soil
- 413 Pollut. 197, 37-47.
- 414 Folk, R.L., 1980. Petrology of sedimentary rocks. Austin Texas: Hemphill Publications. p.182.
- 415 Garriz, A., Pamela, S., Carriquiriborde, P., Miranda, L.A., 2019. Effects of heavy metals
- 416 identified in chascomús shallow lake on the endocrine-reproductive axis of pejerrey fish
- 417 (Odontesthes bonariensis). Gen. Comp. Endocrinol. 273, 152-162.
- Goher, M.E., Farhat, H.I., Abdo, M.H., Salem, S.G., 2014. Metal pollution assessment in the
- 419 surface sediment of lake Nasser, Egypt. Egyptian J. Aquatic Res. 40(3), 213-224.
- 420 Hamdoun, H., Van-Veen, E., Basset, B., Lemoine, M., Coggan, J., Leleyter, L., Baraud., F., 2015.
- 421 Characterization of harbor sediments from the English Channel: assessment of heavy metal
- 422 enrichment, biological effect and mobility. Mar. Pollut. Bull. 90, 273-280.
- 423 IIE, 1998. Boletín IIE mayo-junio/1998. Instituto de Investigaciones Eléctricas. From
- 424 http://iie.org.mx/publica/bolmj98/secmj98.htm. 28 Nov 2019. (In Spanish)
- 425 INEGI, 2018. Anuario Estadístico del Estado de Campeche. Instituto Nacional de Estadística,
- 426 Geografía e Informática. (In Spanish)

- 427 Karar, S., Hazra, S., Das, S., 2019. Assessment of the heavy metal accumulation in the blue
- 428 Swimmer Crab (Portunus pelagicus), northern Bay of Bengal: Role of salinity. Mar. Pollut. Bull.
- 429 143, 101-108.
- 430 Kulkarni, R., Deobagkar, D., Zinjarde, S., 2018. Metals in mangrove ecosystems and associated
- 431 biota: a global perspective. Ecotoxicol. Environ. Saf. 153, 215-228.
- Long, E., MacDonald, D., Smith, S., Calder, F., 1995. Incidence of adverse biological effects
- within ranges of chemical concentrations in marine and estuarine sediments. Environ. Manage.
- 434 19, 81-87.
- Lovelock, C.E., Cahoon, D.R., Friess, D.A., Guntenspergen, G, R., Krauss, K.W., Rogers, K.,
- 436 Saunders, M.L., Sidik, F., Swales, A., Saintilan, N., Thuyen, L.X., Triet, T., 2015. The vulnerability
- of Indo-Pacific mangrove forests to sea-level rise. Nature. 526, 559-563.
- 438 Mafi-Gholami, D., Zenner, E. K., Jaafari, A., and Ward, R. D., 2019. Modelling multi-decadal
- 439 mangrove leaf area index in response to drought along the semi-arid southern coasts of Iran.
- 440 Sci. Total Environ. 656, 1326-1336.
- 441 Mehaur, S., Filatova, T., Dastgheib, A., de Ruyter van Steveninck, E., Ranasinghe, R., 2018.
- 442 Quantifying economic value of coastal ecosystem services: a review. J. Mar. Sci. Eng. 6(1), 5.
- 443 McLeannan S. M., 2001. Relationship between the trace element composition of sedimentary
- rocks and upper continental crust. Geochem. Geophys. Geosyst. Doi: 10.1029/200G000109.
- Müller, G., 1969. Index of geoaccumulation in sediments of the Rhine River. Geol. J. 2, 109-118.
- 446 Muños, A., Gutierrez, E.A., Daesslé, M.V., Orozco, J.A., 2012. Relationship between metal
- 447 enrichments and a biological adverse effects index in sediments from Todos Santos Bay,
- northwest coast of Baja California, Mexico. Mar. Pollut. Bull. 72, 6-13.

- Nava, J.C., Arenas, P., Cardoso, F., 2018. Integrated coastal management in Campeche, Mexico:
- a review after the Mexican marine and coastal national policy. Ocean Coast. Manage. 154, 34-
- 451 45.
- 452 NOAA, 2014. Oil Spills in Mangroves. National Oceanic and Atmospheric Administration
- 453 http://response.restoration.noaa.gov/sites/default/Oil\_Spill\_Mangrove.pdf, Acceded date:
- 454 November 2019.
- 455 NOAA, 2016. Screening Quick Reference Tables. National Oceanic and Atmospheric
- 456 Administration http://response.restoration.noaa.gov/sites/default/files/SQuiRTs.pdf, Acceded
- 457 date: November 2019.
- 458 Ortiz, L.E., Escamilla, J.S., Flores, K., Ramirez, M., Acevedo, O., 2006. Características geológicas
- 459 y potencial metalogenético de los principales complejos ultramaficos- Maficos de México.
- 460 Boletín de la Sociedad Geológica Mexicana Tomo LVII. 4: 161-181. (in spanish)
- 461 PEMEX 2018. Anuario Estadístico (in Spanish). Petróleos Mexicanos.
- 462 Ramos-Vázquez, M.A., Armstrong-Altrin, J.S., 2019. Sediment chemistry and detrital zircon
- record in the Bosque and Paseo del Mar coastal areas from the southwestern Gulf of Mexico.
- 464 Mar. Petrol. Geol. 110, 650-675.
- Ruiz, A.C., Sprovieri, M., Piazza, R., Frignani, M., Sanchez, J.A., Alonso, C., Martinez, U., Perez,
- 466 L.H., Preda, M., Hillaire, C., Gastaud, J., Quejido, A.J., 2012. Effects of land use change and
- 467 sediment mobilization on coastal contamination (Coatzacoalcos River, Mexico). Cont. Shelf
- 468 Res. 37, 57-65.
- 469 Ruiz, A.C., Sanchez, J.A., Perez, L.H., Gracía, A., 2019. Spatial and temporal distribution of
- 470 heavy metal concentrations and enrichment in the southern Gulf of Mexico. Sci. Total Environ.
- 471 651, 3174-3186.

- 472 Senthilkumar, B., Purvaja, R., Ramesh, R., 2013. Vertical distribution and accumulation of
- heavy metals in mangrove sediments (Pichavaram), southeast coast of India. J. Appl. Geochem.
- 474 15, 318-335.
- 475 Shepard, F.P., 1954. Nomenclature based on sand-silt-clay ratios. J. Sediment. Petrol. 24, 154-
- 476 158.
- 477 Shi, C., Ding, H., Zan, Q., Li, R., 2019. Spatial variation and ecological risk assessment of heavy
- 478 metals in mangrove sediments across China. Mar. Pollut. Bull. 143, 115-124.
- 479 Siddique, M.A.A., Mustafa Kamal, A.H., Aktar, M., 2012. Trace metal concentrations in salt
- 480 marsh sediments from Bakkhali River estuary, Cox's Bazar, Bangladesh. Zoo. Ecol. 22, 254-259.
- 481 Singh, U.K., Kumar, B., 2017. Pathways of metals contamination and associated human health
- 482 risk in Ajay River basin, India. Chemosphere 174, 183-199.
- 483 Souza, I., Rocha, L., Morozesk, M., Bonomo, M., Arrivabene, H., Duarte, I., Furlan, L.,
- 484 Monferrán, M., Mazik, K., Elliott, M., Matsumoto, S., Milanez, C., Wunderlin, D. and Fernandes,
- 485 M., 2015. Changes in bioaccumulation and translocation patterns between root and leafs of
- 486 Avicennia schaueriana as adaptive response to different levels of metals in mangrove system.
- 487 Mar. Pollut. Bull. 94, 176-184.
- Souza, I., Arrivabene, H., Craig, C., Midwood, A., Thornton, B., Matsumoto, S., Elliott, M.,
- 489 Wunderlin, D., Monferrán, M. and Fernandes, M., 2018. Interrogating pollution sources in a
- 490 mangrove food web using multiple stable isotopes. Sci. Total Environ. 640, 501-511.
- Tomlinson, D.L., Wilson, J.G., Harris, C.R., Jeffrey, D.W., 1980. Problems in the assessment of
- 492 heavy metal level in estuaries and the formation of pollution index. Helgol. Mar. Res. 33, 566-
- 493 575.

- Veetil, B. K., Ward, R. D., Quang, N., Trang, N., and Giang, N.T.T., 2019. Mangroves of Vietnam:
- 495 Historical development, current state of research and future threats. Estuar. Coast. Shelf S.
- 496 218, 212-236.
- 497 Ward, R., Friess, D., Day, R., and Mackenzie, R., 2016. Impacts of climate change on global
- 498 mangrove ecosystems: a regional comparison. Ecosyst. Health Sustain. 2(4), 1-25.
- 499 Zhang, Z., Wang, J.J., Tang, C., Delaune, R.D., 2015. Heavy metals and metalloids content and
- enrichment in gulf coast sediments in the vicinity of an oil refinery. J. Geochem. Explor. 159,
- 501 93-100.



\*Highlights (for review : 3 to 5 bullet points (maximum 85 characters including spaces per bullet point)

# Highlights

- City sewages and boatyards act as point sources for Zn, Cu and Pb.
- Ni and V contents show that oil industry has a low impact in mangroves.
- Zn, Cu, and Pb contents exceed adverse effect index (AEI) in two urban point sources.

- 1 Environmental risk of trace elements in mangrove ecosystems: An assessment of
- 2 natural vs oil and urban inputs
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- 4 Canales-Delgadillo Julio César<sup>ab</sup>; Pérez-Ceballos Rosela Yazmin<sup>ab</sup>; Ward, Raymond D.<sup>e,f</sup>;
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## Highlights

- City sewages and boatyards act as point sources for Zn, Cu and Pb.
- Ni and V contents show that oil industry has a low impact in mangroves.
- Zn, Cu, and Pb contents exceed adverse effect index (AEI) in two urban point sources.

#### Abstract

- The petrochemical industry and urban activities are widely recognized worldwide as a source of pollution to mangrove environments. They can supply pollutants such as trace elements that can modify the ecosystem structure and associated services, as well as human populations. Through geochemical data, multivariate statistical analysis and pollution indices such as the enrichment factor (EF), geo-accumulation index (Igeo), adverse effect index (AEI) and the pollution load index (PLI), we evaluated the factors that control trace element distribution, punctual sources and determined the pollution level of sediments and their potential biological impact in the mangrove ecosystem of Isla del Carmen, Mexico. The factor and cluster analysis highlighted that the distribution of trace elements is influenced by the mineralogy, texture as well as urban derived sources. The pollution indices showed values in the punctual sources from the urban area of EF > 10, Igeo > 3, AEI >3, PLI > 1 by Cu, Zn and Pb. Finally, the results revealed that mangroves from Isla del Carmen has a major influence from urban activities and natural sources rather than oil industry and also indicate a degraded environment as a result of anthropogenic activities that could have knock-on effect for human health if polluted marine organisms derived from the urban mangroves are consumed.
- **Capsule abstract:** Surface sediments show the influence of point sources on selected trace element concentrations correlated with human activities within the mangroves of Isla del Carmen, Mexico.
- **Keywords:** Trace Elements, Sediment, Oil Industry, Mangroves, Gulf of Mexico, Isla del Carmen

#### 1. Introduction

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Mangroves are vegetated intertidal coastal environments located in tropical and subtropical areas. They are characterized by the presence of plants adapted to frequent inundation by saline water imposed by a tidal regime (NOAA, 2014). To support and sustain the well-being of human being, mangroves provide a range of ecosystem services such as: nursery and breeding habitats for commercial shrimp, crabs and fish species; coastal protection against erosion processes, cyclones, storms, floods and carbon sequestration (Brander et al., 2012; Lovelock et al., 2015; Mehaur et al., 2018; Shi et al., 2019). Like many other ecosystems, mangroves have been subjected to anthropogenic pressures resulting in decreases in mangrove area, particularly conversion to agriculture, aquaculture and salt ponds, urban expansion as well as alterations to sediment supply, sea level rise, drought and storms as a result of climate change (Ward et al., 2016; Veetil et al., 2018; Mafi-Gholami et al., 2019). Due to their location in coastal and estuarine settings they are also often exposed to a range of pollutants particularly those located proximal to industry or urban settings. Many of these pollutants are harmful to mangrove associate organisms and humans when transferred through the trophic chain (Senthilkumar et al., 2013; Kulkarni et al., 2018; Chai et al., 2019). Trace elements are common pollutants found in mangroves and are considered as a serious threat to human health and other living organisms due to their persistence, potential for bioaccumulation, and toxicity (Arrivabene et al., 2016; Celis et al., 2017; Singh and Kumar, 2017). However, it is important to note that trace elements can be derived from anthropogenic and natural sources that complicate analysis concerning their environmental impact (Anaya-Gregorio et al., 2018; Armstrong-Altrin et al., 2019). To resolve this issue geochemical indices such as the enrichment factor (EF), geoaccumulation Index (Igeo), pollution load index (PLI), adverse effect index (AEI), as well as statistical tools have been developed to provide a better understanding about the environmental risks that these

- substances pose to mangrove ecosystems and humans (Goher et al., 2014; Duodu et al., 2017;
- 73 Benson et al., 2018).
- 74 Mangroves from Isla del Carmen are likely to be influenced by point source pollution from
- 75 urban, industrial and petrochemical derived sources as a result of the proximity to an industrial
- 76 port city and petrochemical source. Therefore, Isla del Carmen is an important area to study
- how oil industry and urban activities impact mangrove environments.
- 78 The aim of this study is to determine the agents that influence trace element distribution in
- 79 mangroves with pollution sources including the oil industry and urban activities. Additionally,
- 80 we aimed to evaluate the contamination level of sediments and their potential biological
- 81 impact in the mangroves of Isla del Carmen, Mexico.

# 2. Study area

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Isla del Carmen is located in the southeastern Gulf of Mexico in Campeche, Mexico (Figure 1). Twenty three percent of its surface corresponds to the urban area of Ciudad del Carmen, while the remaining 77% is covered by mangroves (INEGI, 2018). Ciudad del Carmen is the second most populated urban center (248,303 inhabitants) in Campeche state (INEGI, 2018). Isla del Carmen belongs to a natural protected area known as Wildlife Protection Area Laguna de Terminos, which includes the largest coastal lagoon environment in Mexico (about 7050.16 Km²) (INEGI, 2018). In the north coast of the island the marine environment dominates and houses the Canterell oil field, which is the largest gas (38%) and offshore oil (56.5%) producer in Mexico (Nava et al., 2018; PEMEX, 2018) while, in the south coast it receives fluvial discharges of the Palizada, Chumpan and Candelaria rivers, which supply nutrients, sediments, as well as pollutants generated by agricultural, cattle feeding and other human activities (Carvalho et al., 2009). The weather is mostly humid warm with summer rainfall. The average annual temperature ranges from 26 to 28 °C, while the average annual

rainfall ranges from 1500 to 2000 mm (INEGI, 2018). Seasonally, three different regimes are recognizable, the dry season (Feb-May), rainy season (Jun-Sep), and stormy season (Oct-Jan).

## 3. Materials and methods

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Sediment samples were collected from thirty-six mangrove sites on Isla de Carmen, Mexico in August 2019 that corresponded to the rainy season; this season was chosen because the rivers' influence is higher in the area rather than other seasons (Figure 1). Samples were collected from creek edges directly adjacent to the mangroves using a Van Veen dredge from a boat at a water column depth from 0.5 to 2.5 m at mid-high tide to avoid running aground. In order to avoid contamination during sampling sediment samples were removed with a plastic spatula from the middle part of the dredge in such a way that sediments had no contact with the metallic parts of the equipment and to get the most recently deposited sediment, we collected the samples in first centimetre. Each sample was stored in sample bags at 4°C until chemical analysis. Samples were divided in two halves with one half used for geochemical analysis and the other for granulometry. Samples were dried at 50 °C for 5 days and homogenized in an agate mortar by hand. All concentrations in this work were expressed with reference to dry weight. Major and trace element concentrations were determined using a RIGAKU ZSX Primus II X-ray fluorescence spectrometer system and analyzed in pressed powder briguettes. The accuracy of major elements was evaluated using the rock standard Argillite whose values ranged from 102.6 to 93%, with the exception of  $Na_2O$  (70.0%),  $P_2O_5$  (73.1%) and  $K_2O$  (89.3%). The accuracy of trace elements was estimated using the standard CH-1 marine sediment, values ranged from 102.6 to 90.8% with the exception of V (111.8%) and Co (113.8%). Sediment particle size distribution was determined by sieving with a RX-29 Ro-Tap Sieve Shaker and a standard ASTM sieve set from -2.0, -1.0, 0.0, 1.0, 2.0, 3.0, and 4.0 Φ. Sediments were classified as mud, sand, and gravel

based on the methodology proposed by Folk (1980). Organic matter and carbonates were

estimated using a loss on ignition method: 1 g of sediment samples were weighed before and after incineration in a muffle furnace at 550 °C and again at 950 °C for 24 hrs in order to eliminate the organic matter and carbonates (Ahmed et al., 2018). Total organic matter and carbonates are reported as percentage of dry weight.

The results were interpreted using descriptive and multivariate statistical tools including correlation analysis to evaluate the strength of relationship among geochemical parameters. A cluster analysis was used to identify impacted locations in mangroves from Isla del Carmen, and factor analysis was performed to determine the variables that better explain the system variability (Statistical analyses were performed using the software STATISTIC 8). Geochemical indices such as EF, Igeo, PLI, and AEI were employed to evaluate the natural or anthropogenic origin for each trace element: to compare and contrast environmental quality with other locations inside the island, and to infer any possible adverse effect on benthic biota (Celis et al., 2018).

**Enrichment Factor** 

135 This proxy is evaluated using equation (1)

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$$EF = \frac{\frac{X_{\text{sample}}}{Y_{\text{sample}}}}{\frac{X_{\text{background}}}{Y_{\text{background}}}}.....(1)$$

Where  $X_{sample}$  and  $X_{background}$  are the concentrations of any trace element analysed in the sample and  $Y_{sample}$  and  $Y_{background}$  are the concentrations of a conservative element that is used to normalize the data. In this work, we used  $Al_2O_3$  as a conservative element because it represents the clay fraction in the sediment and it is not affected by anthropogenic sources or geochemical changes such as redox processes (Celis et al., 2013). As background we employed the least contaminated station (Feng et al., 2010) rather than the Upper Continental Crust (UCC) (McLennan, 2001) to avoid any overestimation in the data interpretation. The

- enrichment factor results were divided into ranks to establish the degree of enrichment (Birth,
- 145 2003) (Table S2).
- 146 Geo-accumulation index
- 147 This proxy is evaluated using equation 2 (Müller, 1969).

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$$I_{geo} = log_2 \left( \frac{C_n}{1.5B_n} \right)$$
 (2)

- 149 Where "C<sub>n</sub>" represents the concentration of any trace element in the sediment sample and
- 150 "B<sub>n</sub>" represents the background concentration of the same element; the value of 1.5 is a factor
- that considers possible variability generated by lithological changes. Geo-accumulation index
- results were divided into ranks to establish the degree of pollution (Birth, 2003) (Table S2).
- 153 Adverse Effect Index

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- The adverse effect index (AEI) can be applied to evaluate how many times the metal concentration in the sediment has exceeded the Threshold Effect Level (TEL) developed by Long et al., (1995); as well as inferring whether trace element concentrations in the sediment could adverse impact benthic biota (Muños et al., 2012; Hamdoun et al., 2015; Baptista et al., 2017). An AEI lower than 1 means that the trace element concentration in the samples is not high enough to produce adverse effects from sediment contamination to organisms where no
- 161 This proxy is calculated using equation (3)

toxicity studies are undertaken or available.

$$AEI = \frac{[MC]}{[SQG_s]}$$

Where MC is the trace element concentration in the sample and SQGs is any sediment quality guideline such as Effect Range Low (ERL), Threshold Effect Level (TEL) and others. SQGs are based on the collection, revision and the integration of a range of studies performed in North

America using laboratory bioassays, equilibrium-partitioning modelling and field studies on the toxicity of metals in sediment on benthic composition (Long et al., 1995). In this work, we used the Threshold Effect Level (TEL) and the Analogous Effect Threshold (AET). Both were obtained via the Screening Quick Reference Tables (SQuirRts) supplied by NOAA, 2016.

Pollution Load Index

The pollution load index is calculated using equations 4 and 5 (Tomlinson et al., 1980).

$$PLI_{sample} = \sqrt[n]{CF_{metal 1} * CF_{metal 2} * ... CF_{metal n}} ... ... ... ... ... (4)$$

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$$CF = \left[\frac{C_{mS}}{C_{mB}}\right]$$
.....(5)

In general; the PLI is calculated with the  $n^{th}$  root of the variable called contamination factor (CF) which is evaluated using the equation (4). The super index "n" in the root represents the total number of samples. While the CF is defined as the ratio between the element concentration analyzed in the sample " $C_{ms}$ " and the background concentration of the same element that is identified in the equation (5) as " $C_{mB}$ ". The background was derived using the least contaminated station (Feng et al., 2010). According to Tomlinson et al., (1980), PLI values of zero suggest the absence of pollutants, while PLI values of one or greater denotes the presence of pollutants or progressive deterioration of sediment quality.

#### 4. Results

# 4.1. Textural analysis

Mud (silt and clay) content in the samples varied from 26.5 to 92.1%; while sand and gravel varied from 6.0 to 67.0% and 0.05 to 40.3%, respectively (Table 1). The mud, sand and gravel ternary diagram (Figure 2) showed that 75.1% of the samples were classified as mud, 13.8% were classified as muddy-sand and 11.1% were not well classified because they had similar values for each of these parameters.

## 4.2. Geochemical composition

The *average* organic matter (OM) content in the study area was 18.1% and varied between 6.8 and 36.4%. The highest concentration was found at station 6, which was located within the urban area of Ciudad del Carmen (Figure 1). The lowest concentration was found at station 31, approximately 35 km from station 6, such a contrast suggests that organic carbon could be associated to the local drainage piping of Ciudad del Carmen. The *average* carbonate concentration was 21.2%. The highest carbonate values were found in the middle of the island at stations 25 (32.4%), 26 (37.5%) and 27 (33.9%), which are characterized by the presence of shell banks and higher energy conditions. The lowest values where found at station 11 (8.3%) and 21 (6.5%). Major element concentrations varied within the following ranges: 11.4 – 35.3% for CaO, 11.2 – 34.5% for SiO<sub>2</sub>, 2.23 – 9.53% for Al<sub>2</sub>O<sub>3</sub> and 1.49 – 4.52% for Fe<sub>2</sub>O<sub>3</sub> (Table 1). The *range* and *average* of trace elements are shown in Table 1. Stations 7 and 1 that are located close to the city sewages exhibited the highest values for Zn (794 mg kg<sup>-1</sup>, 365 mg kg<sup>-1</sup>), Cu (311 mg kg<sup>-1</sup>, 81mg kg<sup>-1</sup>) and Pb (111 mg kg<sup>-1</sup>, 48 mg kg<sup>-1</sup>) (Figure 3, Figure S2). Station 10 situated close to a boat yard, where there is limited water flux because of mangroves exhibited high values of Zn (471 mg kg<sup>-1</sup>), Cu (35 mg kg<sup>-1</sup>) and Pb (34 mg kg<sup>-1</sup>) (Figure 3, Figure S2).

#### 4.3. Pollution indices

# 4.3.1. Enrichment Factor

All the trace elements showed variations in their enrichment level degrees (Figure 4). All samples exhibited levels from no enrichment to minor enrichment for V, Cr, Rb, Sr, Ba and Zr. Moderate to severe EFs were found for Co and Ni at station 30 (5.1) and 30 (5.3) respectively. Values of EFs > 10 (severe and very severe enrichment) were found for Cu at stations 1 (24.5) and 7 (63.1), for Zn at stations 1 (27.6), 7 (40.3) and 10 (12.2), and for Pb at stations 1 (14.5) and 7 (22.5).

#### 4.3.2. Geo-accumulation Index

The geo-accumulation Index shows (Figure 4) that none of the samples were moderately polluted with Sr, Ba and V. Between 4 and 14 of the samples were moderately polluted with Cr, Zr, Zn, Ni, Rb, Cu, Pb and Co (Figure 4). The highest values of Igeo were found at stations 1 (4.2, 6.3, 3.4), 7 (6.1, 7.5, 4.6), 10 (3, 6.7, 2.9), 29 (3.4, 4.7, 1.7) and 30 (1.6, 3.7, 1.2), being moderately to extremely polluted with Cu, Zn and Pb, respectively.

#### 4.3.3. Adverse Effects Index

The estimated AEI values for the mangroves of Isla del Carmen showed that the samples could be associated with adverse effects on organisms caused by Cr (100%), Ba (100%), Ni (47.2%), Co (25%), Cu (11.1%), V (8.3%) and Zn (8.3%), (Figure 4). The *maximum* AEI values were found at station 7 for Cu (16.6), Zn (6.4), Pb (3.7) and Ba (3.2). While the *maximum* values of AEI at station 11 corresponded to V (1.2), and Ni (5.1), at station 9 the highest AEI corresponds to Co (1.9) and they are the at stations located within a small boat channel next to the urban mangroves.

#### 4.3.4. Pollution Load Index

Calculated PLI values of > 1 were recorded at every site (Figure 5), which means that 100% of the samples had a significant deterioration of sediment quality due to trace element concentrations. However, for this index, there was an evident tendency of samples collected inside the tidal channels, and around Ciudad del Carmen, to exhibit PLI values > 2 (except station 13). Meanwhile, samples collected at places far away from human activity showed PLI values < 2 (except stations 29 and 30).

#### 4.4. Factor and cluster analysis of sediment samples

A factor analysis was performed to identify the causes of variation between the geochemical and textural data from the mangroves. Four factors were identified that explained 81% of the system's *variance*. All factors were associated with the mineralogy, texture, anthropogenic and

natural sources (Table 2). The first factor explained 29% of the total *variance* and was related with sediment mineralogy due to its interactions between major elements, carbonates and gravel. The second factor explained 20% of the total *variance*, and it was related to sediment texture because this factor grouped parameters such as sand, mud, OM, CaO, Na<sub>2</sub>O, and S, which are widely associated to a specific sediment grain size. The third factor explained 17% of the total *variance* and they grouped P<sub>2</sub>O<sub>5</sub>, Cu, Zn, and Pb. It was linked to anthropogenic sources because the pollution indices noted that these elements are supplied by anthropogenic and point sources. Finally, the fourth factor explained 15% of the total *variance*; it grouped V, Co, Ni and Rb, which was linked to natural sources because according to the pollution indices these elements were likely to be supplied by the Palizada, Chumpan and Candelaria river catchments lithology.

In order to identify associations between 36 sediment samples, a cluster analysis was performed. The cluster diagram showed three main groups (A, B and C) (Figure 6). Group A was characterized by a higher content of gravel, sand, OM and CO<sub>3</sub>. Group B was characterised by higher concentrations of SiO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>, K<sub>2</sub>O, MgO, V, Cr, Ni, Rb, Zr. Group C was characterised by high mud content, OM, Fe<sub>2</sub>O<sub>3</sub>, P<sub>2</sub>O<sub>3</sub>, S, Co, Cu, Zn, Pb, and Ba. Group C also showed the highest PLI values and highest EF, Igeo and AEI of Co, Cu, Zn, Pb and Ba.

## 5. Discussion

## 5.1. Major and trace elements source

The geochemical composition and the ternary diagram (Figure S1) showed that mangrove sediments in the area are richer in CaO (Carbonates) than  $Al_2O_3$  (aluminosilicates) and  $Fe_2O_3$  (heavy minerals). Carbonates are supplied by shell fragments of organisms that live in the mangrove such as snails, oysters and other mollusks and bivalves. The negative *correlation* between CaO and CO<sub>3</sub> (0.48, p < 0.05) suggests that although the study area is characterized by biogenic sediment rich in carbonates, there are other calcium sources such as plagioclases,

otherwise this *correlation* would be high. The negative *correlation* between CaO and MgO (-0.74, p < 0.05) noted that Mg substitutes Ca from carbonate minerals due to its similar radio atomic size. The positive *correlation* between Fe<sub>2</sub>O<sub>3</sub> and TiO<sub>2</sub> (0.95, p < 0.05) is related to heavy minerals. Positive *correlation* between Al<sub>2</sub>O<sub>3</sub> with SiO<sub>2</sub> (0.90, p < 0.05), and K<sub>2</sub>O (0.93, p < 0.05), highlighted the presence of aluminosilicates such as plagioclases and the positive *correlation* between Al<sub>2</sub>O<sub>3</sub> with SiO<sub>2</sub> (0.90, p < 0.05), Fe<sub>2</sub>O<sub>3</sub> (0.92, p < 0.05), TiO<sub>2</sub> (0.89, p < 0.05), and K<sub>2</sub>O (0.93, p < 0.05) suggests that all of these minerals are in fine particles such as mud rather than sand and gravel. On the other hand, negative *correlations* were recorded between CO<sub>3</sub> and SiO<sub>2</sub> (-0.88, p < 0.05), Al<sub>2</sub>O<sub>3</sub> (-0.82, p < 0.05), Fe<sub>2</sub>O<sub>3</sub> (-0.75, p < 0.05), TiO<sub>2</sub> (-0.75, p < 0.05), K<sub>2</sub>O (-0.88, p < 0.05), indicating that biogenic sourced sediments were diluted by terrigenous sediment from the Palizada, Chumpan and Candelaria rivers and it is supported by the ternary diagram (Figure S1) where it is clear how CaO concentration decreases when sediment is getting rich in Al<sub>2</sub>O<sub>3</sub>.

The trace element data and the cluster analysis (Figure 6) highlighted stations 7, 10 and 1 as punctual sources, in these stations the highest trace elements concentration and pollution indices of Zn, Cu and Pb are identified. The lack of positive *correlation* for these elements with almost all the studied parameters and their spatial distribution support the idea that the city sewages and boat yards act as anthropogenic sources for these elements (Table S1). In contrast, the positive *correlations* for V (0.59, 0.61, 0.57, 0.87, p<0.05), Ni (0.65, 0.60, 0.56, 0.86, p<0.05), with Al<sub>2</sub>O<sub>3</sub>, Fe<sub>2</sub>O<sub>3</sub>, TiO<sub>2</sub> and Co suggest a terrestrial natural origin probably supplied by Palizada, Chumpan and Candelaria rivers, where their lithology are linked to mafic rocks (Ortiz et al., 2006). Relatively low *correlations* between Ba and Zr with almost all major and trace elements (Table S1) suggest an oil industry origin; as these elements are likely to be related to crude oil, residual fuel oils, accidental oil spills and oil drilling (Fieldler et al., 2009; Zhang et al., 2015). On the other hand, the cluster analysis showed that the association between stations were divided by three geochemical groupings according to their mineralogy,

texture and anthropogenic/natural influence. This is supported by the factor analysis results where the mineralogy, texture, anthropogenic and natural sources were identified as factors that explained the geochemical variations from the mangrove sediments of Isla del Carmen.

### 5.2. Evidence of oil influence in the Carmen Island Mangroves

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V and Ni are elements that are related to oil industry because they are abundant in crude oil. Generally, the ratio Ni/V is used to identify oil families such as Brent, West Texas, Maya, etc (Barwaise, 1990). However, this ratio also has been used as an oil pollution index in places with presence of an important oil industry such as the Gulf of Mexico (Cuevas et al., 2018; Ruiz et al., 2019). We used this ratio to identify the oil industry influence on the mangroves of Isla del Carmen. According to Barwaise (1990) the content of Ni and V in crude oil is about 340 mg kg<sup>-1</sup> and 1580 mg kg<sup>-1</sup> respectively. Also, according to the IIE (1998), the chemical composition of the Mexican fuel oil could have up to 20-60 and 200-350 mg L<sup>-1</sup> of Ni and V respectively, giving two interval ratios from 0.3 to 0.17 and 0.1 to 0.05 evidencing geochemical markers that Ni and V come from Mexican oil industry activities. The ratios found in the mangroves of Isla del Carmen ranged from 0.5 to 1.5. Such values are higher than the values reported average in Veracruz (0.28) and Tamaulipas (0.17 - 0.23), states with high oil industry activity (Ruiz et al., 2012; Celis et al., 2018). These results suggest a low impact from oil industry sources and confirm the predominantly natural source of these trace elements. Recent studies on coastal sediments near to the Isla del Carmen also reported that the Ni and V contents are due to the natural sources and were derived by the contribution of intermediate and mafic source rocks like andesite and basalt (Armstrong-Altrin et al., 2018; Ramos-Vázquez and Armstrong-Altrin, 2019).

### 5.3. Ecological status of the area

The pollution indices such as the PLI reported in this study suggest that the Isla del Carmen mangroves are considered a degraded environment as a result of the trace elements present

in the sediments. The EF, Igeo and AEI highlighted specific sites close to point sources that caused severe enrichment, very severe enrichment and those sites were classified as strongly polluted and extremely polluted mangrove environments for Zn, Cu and Pb.

The adverse effects that are likely to be evident in the mangrove ecosystem as a result of these trace element concentrations would be bioaccumulation and biomagnification in species at all the trophic levels (Baki et al., 2018; Karar et al., 2019). In addition, because of their toxicity, mangrove organisms would show a reduction in their growth development, as well as sexual and genetic anomalies (Souza et al., 2015), and even there would be a decrease in the number of benthic species (crabs, clams), fish and birds (Souza et al., 2018; Garriz et al., 2019) that coexist in the mangrove environment. Also, it could affect other proximal coastal environments such as the Terminos Lagoon, and the estuarine environment that exists in this natural system. Finally, it could affect human health if polluted marine organisms derived from the mangroves are consumed by the human population; as these levels of the trace elements are likely to cause serious damage to the kidney, central nervous system, liver, intestinal tract and reproductive system (Siddique et al., 2012; Abdel-Khalek et al., 2016). However, it is necessary to conduct further investigation to assess the impacts of these pollutants on local ecological diversity and biodiversity to fully assess the ecological status in the area.

## 6. Conclusion

The bulk geochemical data noted that almost all the trace element concentrations in the mangroves from Isla del Carmen are close to background levels; except for Zn, Cu and Pb where the spatial distribution suggests that the city sewages and boatyards act as point sources for these elements. The trace element distributions in the mangroves from Carmen Island are driven by natural and anthropogenic factors such as the mineralogy, texture and urban activities or urban sources. Finally, the mangroves from the Carmen Island can be considered impacted by the wastes generated within the urban area rather than by the

extraction of oil activities that take place close to the Carmen Island, which suggests that there is limited impact from oil activities in the local area. However, levels are such that these areas can be considered as heavily polluted by some of the trace elements investigated and this is likely to cause adverse effects on benthic organisms associated with the mangroves.

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### References

- 350 Abdel-Khalek, A.A.A., Elhaddad E., Mamdouh S., Assm, M.S.M., 2016. Assessment of Metal
- 351 Pollution around Sabal Drainage in River Nile and its Impacts on Bioaccumulation Level, Metals
- 352 Correlation and Human Risk Hazard using Oreochromis niloticus as a Bioindicator. Turk. J. Fish.
- 353 Aquat. Sci. 16, 227-239.
- 354 Ahmed, I., Mostefa, B., Bernard, A., Oliver, R., 2018. Levels and ecological risk assessment of
- 355 heavy metals in surface sediments of fishing grounds along Algerian coast. Mar. Pollut. Bull.
- 356 136, 322-333.
- 357 Anaya-Gregorio, A., Armstrong-Altrin, J.S., Machain-Castillo, M.L., Montiel-García, P.C., Ramos-
- 358 Vázquez, M.A., 2018. Textural and geochemical characteristics of late Pleistocene to Holocene
- 359 fine-grained deep-sea sediment cores (GM6 and GM7), recovered from southwestern Gulf of
- 360 Mexico. J. Palaeogeog. 7(3), 253-271.

- 361 Armstrong-Altrin, J.S., Ramos-Vázquez, M.A., Zavala-León, A.C., Montiel-García, P.C., 2018.
- 362 Provenance discrimination between Atasta and Alvarado beach sands, western Gulf of Mexico,
- 363 Mexico: Constraints from detrital zircon chemistry and U-Pb geochronology. Geol. J. 53(6),
- 364 2824-2848.
- 365 Armstrong-Altrin, J.S., Botello, A.V., Villanueva, S.F., Soto, L.A., 2019. Geochemistry of surface
- 366 sediments from the northwestern Gulf of Mexico: implications for provenance and heavy
- metal contamination. Geol. Quarter. 63 (3), 522-538.
- 368 Arrivabene, H., Quenupe, C., da Costa, I., Wundorlin, D., Rozindo, D., Rodrigues, S., 2016.
- 369 Differential bioaccumulation and translocation patterns in three mangrove plants
- experimentally exposed to iron: Consequences for environmental sensing. Environ. Pollut. 215,
- 371 302-313.
- Baki, M.A., Hossain, Md., Akter, J., Quraishi, Sh. B., Shohib, Md. F.H., Ullah, A.K.M.A., Khan, Md.
- 373 F., 2018. Concentration of heavy metals in seafood (fishes, shrimp, lobster and crabs) and
- human health assessment in Saint Martin Island, Bangladesh. Ecotoxicol. Environ. Saf. 159,
- 375 153-163.
- Baptista, J.A., Ferreira, C., Gutterres, C., Monteiro, E., Vaz, G., Monica, O., 2017. Environmental
- 377 change in Guanabara Bay, SE Brazil, based in microfaunal, pollen and geochemical proxies in
- 378 sedimentary cores. Ocean Coast. Manag. 143, 4-15.
- 379 Barwise A.J.G., 1990. Role of Nickel and Vanadium in petroleum Classification. Energy & Fuels.
- 380 4, 647-652.
- 381 Benson, N.U., Adedapo, A.E., Fred-Ahmadu, O.H., Williams, A.B., Udosen, E.D., Ayeyuyo, O.O.,
- Olajire, A.A., 2018. New ecological risk indices for evaluating heavy metals contamination in
- aquatic sediment: A case study of the Gulf of Guinea. Reg. Stud. Mar. Sci. 18, 44-56.

- Birth, G., Woodcoffe, C.D., Furness, R.A., 2003. a scheme for assessing human impacts on
- 385 coastal environments using sediments. Coastal GIS 2003. Wollongong University Papers in
- 386 Centre for Marine Policy. 14 (Australia).
- 387 Brander, L.M., Wagtendonk, A.J., Hussain, S.S., McVittie, A., Verburg, P.H., de Groot, R.S.,
- 388 Vander Ploeg, S., 2012. Ecosystem service values for mangroves in Southeast Asia: a meta-
- analysis and value transfer application. Ecosyst. Serv. 1, 62-69,
- 390 Carvalho, F.P., Villeneuve, J.P., Cattini, C., Rendon, J., Mota de Oliveira J., 2009. Pesticide and
- 391 PCB residues in the aquatic ecosystems of Laguna de Terminos, a protected area of the coast
- of Campeche, Mexico. Chemosphere 74, 988-995.
- 393 Celis, O., Rosales, L., Carranza, A., 2013. Heavy metal enrichment in surface sediments from
- the SW Gulf of Mexico. Environ. Monit. Assess. 185, 8891-8907.
- 395 Celis, O., Rosales, L., Cundy, A., Carranza, A., 2017. Sedimentary heavy metal(loid)
- 396 contamination in the Veracruz shelf, Gulf of Mexico: A baseline survey from a rapidly
- developing tropical coast. Mar. Pollut. Bull. 119, 204-213.
- 398 Celis, O., Rosales, L., Cundy, A., Carranza, A., Croudace, I., Hernandez, H., 2018. Historical trace
- 399 element accumulation in marine sediments from the Tamaulipas shelf, Gulf of Mexico: An
- assessment of natural vs anthropogenic inputs. Sci. Total Environ. 622-623, 325-336.
- 401 Chai, M., Li, R., Ding, H., Zan Q., 2019. Occurrence and contamination of heavy metals in urban
- 402 mangroves: A case study in Shenzhen, China. Chemosphere 219, 165-173.
- 403 Cuevas, H., Rosales, L., Marquez, A.Z., Carranza, A., 2018. Environmental assessment of a
- 404 mangrove protected area in the southeast Gulf of Mexico through sediment core analysis.
- 405 Environ. Earth Sci. 77, 73.

- 406 Duodu, G.O., Goonetilleke, A., Ayoko, G., 2017. Potential bioavailability assessment, source
- 407 apportionment and ecological risk of heavy metals in the sediment of Brisbane River Estuary,
- 408 Australia. Mar. Pollut. Bull. 117, 523-531.
- 409 Feng, H., Jiang, H., Gao, W., Weinstern, M., Zhang, Q., Zhang, W., 2010. Metal contamination in
- sediments of western Bohay Bay and adjacent estuaries, China. J. Environ. Manage. 34, 1-13.
- 411 Fiedler, S., Siebe, C., Herre, A., Roth, B., Cram, S., Stahr, K., 2009. Contribution of oil industry
- 412 activities to environmental loads of heavy metals in Tabasco low lands, Mexico. Water Air Soil
- 413 Pollut. 197, 37-47.
- 414 Folk, R.L., 1980. Petrology of sedimentary rocks. Austin Texas: Hemphill Publications. p.182.
- 415 Garriz, A., Pamela, S., Carriquiriborde, P., Miranda, L.A., 2019. Effects of heavy metals
- 416 identified in chascomús shallow lake on the endocrine-reproductive axis of pejerrey fish
- 417 (Odontesthes bonariensis). Gen. Comp. Endocrinol. 273, 152-162.
- Goher, M.E., Farhat, H.I., Abdo, M.H., Salem, S.G., 2014. Metal pollution assessment in the
- 419 surface sediment of lake Nasser, Egypt. Egyptian J. Aquatic Res. 40(3), 213-224.
- 420 Hamdoun, H., Van-Veen, E., Basset, B., Lemoine, M., Coggan, J., Leleyter, L., Baraud., F., 2015.
- 421 Characterization of harbor sediments from the English Channel: assessment of heavy metal
- 422 enrichment, biological effect and mobility. Mar. Pollut. Bull. 90, 273-280.
- 423 IIE, 1998. Boletín IIE mayo-junio/1998. Instituto de Investigaciones Eléctricas. From
- 424 http://iie.org.mx/publica/bolmj98/secmj98.htm. 28 Nov 2019. (In Spanish)
- 425 INEGI, 2018. Anuario Estadístico del Estado de Campeche. Instituto Nacional de Estadística,
- 426 Geografía e Informática. (In Spanish)

- 427 Karar, S., Hazra, S., Das, S., 2019. Assessment of the heavy metal accumulation in the blue
- 428 Swimmer Crab (Portunus pelagicus), northern Bay of Bengal: Role of salinity. Mar. Pollut. Bull.
- 429 143, 101-108.
- 430 Kulkarni, R., Deobagkar, D., Zinjarde, S., 2018. Metals in mangrove ecosystems and associated
- 431 biota: a global perspective. Ecotoxicol. Environ. Saf. 153, 215-228.
- 432 Long, E., MacDonald, D., Smith, S., Calder, F., 1995. Incidence of adverse biological effects
- within ranges of chemical concentrations in marine and estuarine sediments. Environ. Manage.
- 434 19, 81-87.
- Lovelock, C.E., Cahoon, D.R., Friess, D.A., Guntenspergen, G, R., Krauss, K.W., Rogers, K.,
- 436 Saunders, M.L., Sidik, F., Swales, A., Saintilan, N., Thuyen, L.X., Triet, T., 2015. The vulnerability
- of Indo-Pacific mangrove forests to sea-level rise. Nature. 526, 559-563.
- 438 Mafi-Gholami, D., Zenner, E. K., Jaafari, A., and Ward, R. D., 2019. Modelling multi-decadal
- 439 mangrove leaf area index in response to drought along the semi-arid southern coasts of Iran.
- 440 Sci. Total Environ. 656, 1326-1336.
- 441 Mehaur, S., Filatova, T., Dastgheib, A., de Ruyter van Steveninck, E., Ranasinghe, R., 2018.
- 442 Quantifying economic value of coastal ecosystem services: a review. J. Mar. Sci. Eng. 6(1), 5.
- 443 McLeannan S. M., 2001. Relationship between the trace element composition of sedimentary
- rocks and upper continental crust. Geochem. Geophys. Geosyst. Doi: 10.1029/200G000109.
- 445 Müller, G., 1969. Index of geoaccumulation in sediments of the Rhine River. Geol. J. 2, 109-118.
- 446 Muños, A., Gutierrez, E.A., Daesslé, M.V., Orozco, J.A., 2012. Relationship between metal
- 447 enrichments and a biological adverse effects index in sediments from Todos Santos Bay,
- northwest coast of Baja California, Mexico. Mar. Pollut. Bull. 72, 6-13.

- Nava, J.C., Arenas, P., Cardoso, F., 2018. Integrated coastal management in Campeche, Mexico:
- a review after the Mexican marine and coastal national policy. Ocean Coast. Manage. 154, 34-
- 451 45.
- 452 NOAA, 2014. Oil Spills in Mangroves. National Oceanic and Atmospheric Administration
- 453 http://response.restoration.noaa.gov/sites/default/Oil\_Spill\_Mangrove.pdf, Acceded date:
- 454 November 2019.
- 455 NOAA, 2016. Screening Quick Reference Tables. National Oceanic and Atmospheric
- 456 Administration http://response.restoration.noaa.gov/sites/default/files/SQuiRTs.pdf, Acceded
- 457 date: November 2019.
- 458 Ortiz, L.E., Escamilla, J.S., Flores, K., Ramirez, M., Acevedo, O., 2006. Características geológicas
- 459 y potencial metalogenético de los principales complejos ultramaficos- Maficos de México.
- 460 Boletín de la Sociedad Geológica Mexicana Tomo LVII. 4: 161-181. (in spanish)
- 461 PEMEX 2018. Anuario Estadístico (in Spanish). Petróleos Mexicanos.
- 462 Ramos-Vázquez, M.A., Armstrong-Altrin, J.S., 2019. Sediment chemistry and detrital zircon
- record in the Bosque and Paseo del Mar coastal areas from the southwestern Gulf of Mexico.
- 464 Mar. Petrol. Geol. 110, 650-675.
- Ruiz, A.C., Sprovieri, M., Piazza, R., Frignani, M., Sanchez, J.A., Alonso, C., Martinez, U., Perez,
- 466 L.H., Preda, M., Hillaire, C., Gastaud, J., Quejido, A.J., 2012. Effects of land use change and
- 467 sediment mobilization on coastal contamination (Coatzacoalcos River, Mexico). Cont. Shelf
- 468 Res. 37, 57-65.
- 469 Ruiz, A.C., Sanchez, J.A., Perez, L.H., Gracía, A., 2019. Spatial and temporal distribution of
- 470 heavy metal concentrations and enrichment in the southern Gulf of Mexico. Sci. Total Environ.
- 471 651, 3174-3186.

- 472 Senthilkumar, B., Purvaja, R., Ramesh, R., 2013. Vertical distribution and accumulation of
- heavy metals in mangrove sediments (Pichavaram), southeast coast of India. J. Appl. Geochem.
- 474 15, 318-335.
- 475 Shepard, F.P., 1954. Nomenclature based on sand-silt-clay ratios. J. Sediment. Petrol. 24, 154-
- 476 158.
- 477 Shi, C., Ding, H., Zan, Q., Li, R., 2019. Spatial variation and ecological risk assessment of heavy
- 478 metals in mangrove sediments across China. Mar. Pollut. Bull. 143, 115-124.
- 479 Siddique, M.A.A., Mustafa Kamal, A.H., Aktar, M., 2012. Trace metal concentrations in salt
- 480 marsh sediments from Bakkhali River estuary, Cox's Bazar, Bangladesh. Zoo. Ecol. 22, 254-259.
- 481 Singh, U.K., Kumar, B., 2017. Pathways of metals contamination and associated human health
- risk in Ajay River basin, India. Chemosphere 174, 183-199.
- 483 Souza, I., Rocha, L., Morozesk, M., Bonomo, M., Arrivabene, H., Duarte, I., Furlan, L.,
- 484 Monferrán, M., Mazik, K., Elliott, M., Matsumoto, S., Milanez, C., Wunderlin, D. and Fernandes,
- 485 M., 2015. Changes in bioaccumulation and translocation patterns between root and leafs of
- 486 Avicennia schaueriana as adaptive response to different levels of metals in mangrove system.
- 487 Mar. Pollut. Bull. 94, 176-184.
- Souza, I., Arrivabene, H., Craig, C., Midwood, A., Thornton, B., Matsumoto, S., Elliott, M.,
- 489 Wunderlin, D., Monferrán, M. and Fernandes, M., 2018. Interrogating pollution sources in a
- 490 mangrove food web using multiple stable isotopes. Sci. Total Environ. 640, 501-511.
- Tomlinson, D.L., Wilson, J.G., Harris, C.R., Jeffrey, D.W., 1980. Problems in the assessment of
- 492 heavy metal level in estuaries and the formation of pollution index. Helgol. Mar. Res. 33, 566-
- 493 575.

- 494 Veetil, B. K., Ward, R. D., Quang, N., Trang, N., & Giang, N.T.T., 2019. Mangroves of Vietnam:
- 495 Historical development, current state of research and future threats. Estuar. Coast. Shelf S.
- 496 218, 212-236.
- Ward, R., Friess, D., Day, R., and Mackenzie, R., 2016. Impacts of climate change on global
- 498 mangrove ecosystems: a regional comparison. Ecosyst. Health Sustain. 2(4), 1-25.
- 499 Zhang, Z., Wang, J.J., Tang, C., Delaune, R.D., 2015. Heavy metals and metalloids content and
- enrichment in gulf coast sediments in the vicinity of an oil refinery. J. Geochem. Explor. 159,
- 501 93-100.

# **Tables and Figures**

Table 1 Range and average concentration of textural and chemical parameters in the surface sediment samples.

Surface sediment								
	Average	Range		Average	Range			
Gravel (%)	6.80	0.05-40.3	Sr (mg kg <sup>-1</sup> )	897.8	451.0-1397.0			
Sand (%)	26.2	6.00-67.0	Ba (mg kg <sup>-1</sup> )	299.3	222.0-413.0			
Mud (%)	67.0	26.5-92.1	Cr (mg kg <sup>-1</sup> )	152.9	57.0-289.0			
CO₃ (%)	21.2	6.50-37.5	Zr (mg kg <sup>-1</sup> )	95.6	43.0-156.0			
OM (%)	18.1	6.80-36.4	Zn (mg kg <sup>-1</sup> )	88.6	12.0-794.0			
S (%)	0.98	0.4-1.8	Ni (mg kg <sup>-1</sup> )	47.5	15.0-96.0			
CaO (%)	23.3	11.4-35.3	V (mg kg <sup>-1</sup> )	44.1	26.0-68.0			
SiO <sub>2</sub> (%)	22.4	11.2-34.5	Rb (mg kg <sup>-1</sup> )	34.3	15.0-55.0			
$Al_2O_3$ (%)	5.26	2.23-9.53	Cu (mg kg <sup>-1</sup> )	22.6	3.00-311.0			
Na₂O (%)	3.64	1.34-7.45	Pb (mg kg <sup>-1</sup> )	13.8	3.00-111.0			
Fe <sub>2</sub> O <sub>3</sub> (%)	2.90	1.49-4.52	Co (mg kg <sup>-1</sup> )	11.1	4.00-22.0			
MgO (%)	2.44	1.41-3.44						
K₂O (%)	0.98	0.49-1.54						
TiO₂ (%)	0.27	0.11-0.47						
MnO (%)	0.06	0.04-0.11						
P <sub>2</sub> O <sub>5</sub> (%)	0.17	0.05-0.73						

Note: n=36, OM=Organic Matter.

Table 2 Factor Analysis for Mangrove Carmen Island

	Factor 1		Factor 2		Factor 3		Factor 4
Gravel	-0.72	Sand	0.66	P <sub>2</sub> O <sub>5</sub>	-0.94	٧	0.87
CO₃	-0.84	Mud	-0.60	Cu	-0.97	Co	0.89
SiO <sub>2</sub>	0.89	ОМ	-0.92	Zn	-0.93	Ni	0.86
$Al_2O_3$	0.91	CaO	0.86	Pb	-0.98	Rb	0.85
Fe <sub>2</sub> O <sub>3</sub>	0.88	Na₂O	-0.79				
TiO <sub>2</sub>	0.89	S	-0.76				
K <sub>2</sub> O	0.88						
MgO	0.82						
MnO	0.70						
Expl.Var	7.96		5.45		4.61		4.05
Prp.Totl	0.29		0.20		0.17		0.15
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Note: n=36, Varimax raw and p<0.05, OM=Organic Matter.

Figure 1 Study area and sampling location

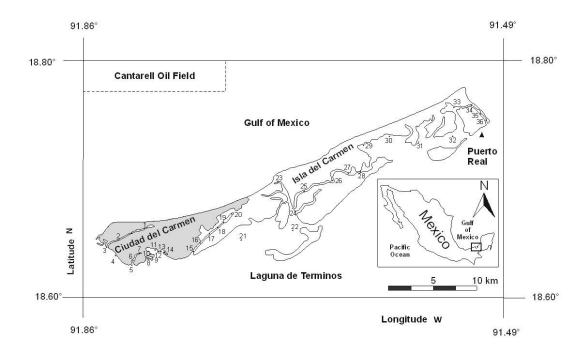


Figure 2 Textural classification of surface sediments after Shepard (1954).

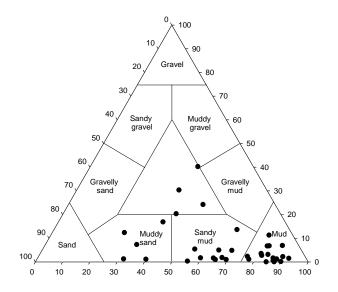


Figure 3 Variation of Ba, Co, Cr, Cu, Ni, Pb, Rb, Sr, V, Zn and Zr. Horizontal lines show the median values over every station.

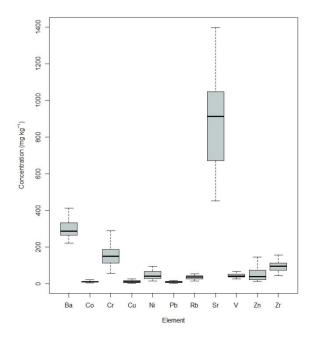


Figure 4 Variations of Enrichment Factors (EFs), Geo-accumulation Index and Adverse Effects Index for Ba, Co, Cr, Cu, Ni, Pb, Rb, Sr, V, Zn and Zr. The colored zones within the plots are gradients of pollutants enrichment: green (0-1) no enrichment, yellow (1-3) minor enrichment, orange (3-5) moderate enrichment, pink (> 5) moderate to severe enrichment.

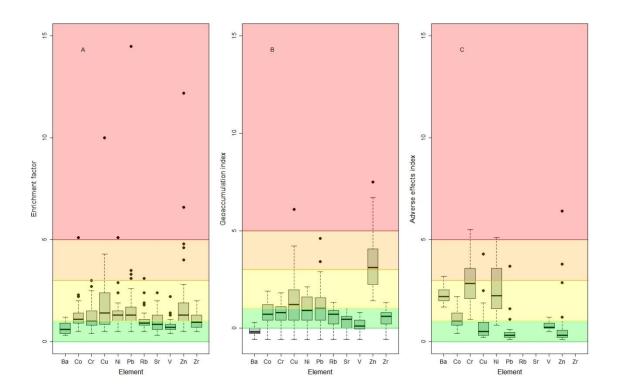
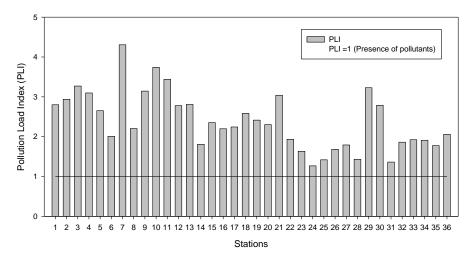
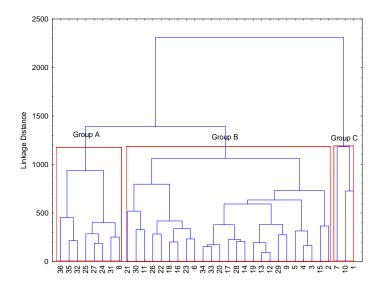


Figure 5 Variation of Pollution Load Index (PLI) for all the stations.



Note: As background the minimum value of the surface sediment sample of each trace element analyzed were used.

Figure 6 Tree diagram from cluster analysis.



Note: n=36 cases, Complete linkage (Manhattan) distances.

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*conflict of Interes	est Statement

We have no competing interests to declare.

### \*Credit Author Statement

### **Credit Author Statement.**

Celis-Hernandez Omar: Conceptualization, Project Administration, Formal Analysis, Writing-Original Draft, Writing-Review and Editing, Resources. Girón-Garcia María Patricia: Investigation, Validation, Resources. Ontiveros-Cuadras Jorge Feliciano: Investigation, Resources. Canales-Canales Delgadillo Julio Cesar: Formal Analysis, Writing-Original Draft. Pérez-Ceballos Rosela Yazmín: Investigation. Ward, Raymond D.: Writing-Original Draft, Writing-Review and Editing. Acevedo-Gonzales Odedt: Investigation. Armsytong-Altring John S.: Writing-Review and Editing, Resources. Merino-Ibarra Martín: Conceptualization. All authors discussed the results and their applications and commented on the manuscript.