

Seed germination in a southern Australian temperate seagrass

Erin Cumming¹, Jessie C. Jarvis^{2,3}, Craig D.H. Sherman¹, Paul H. York^{1,3} and Timothy M. Smith¹

¹ Centre for Integrative Ecology, School of Life and Environmental Sciences, Deakin University, Waurn Ponds, Victoria, Australia

² Department of Biology and Marine Biology, Center for Marine Science, University of North Carolina Wilmington, Wilmington, NC, United States

³ Centre for Tropical Water & Aquatic Ecosystem Research, James Cook University, Cairns, Queensland, Australia

ABSTRACT

In a series of experiments, seeds from a temperate seagrass species, Zostera nigricaulis collected in Port Phillip Bay, Victoria, Australia were exposed to a range of salinities (20 PSU pulse/no pulse, 25 PSU, 30 PSU, 35 PSU), temperatures (13 °C, 17 °C, 22 °C), burial depths (0 cm, 1 cm, 2 cm) and site specific sediment characteristics (fine, medium, coarse) to quantify their impacts on germination rate and maximum overall germination. In southern Australia the seagrass Z. nigricaulis is a common subtidal species; however, little is known about the factors that affect seed germination which is a potential limiting factor in meadow resilience to natural and anthropogenic disturbances. Overall seed germination was low (<20%) with germination decreasing to <10% when seeds were placed in the sediment. When germination of Z. nigricaulis seeds was observed, it was enhanced (greater overall germination and shorter time to germination) when seeds were exposed to a 20 PSU pulse for 24 h, maintained at salinity of 25 PSU, temperatures <13 °C, in sediments with fine or medium grain sand and buried at a depth of <1 cm. These results indicate that germination of Z. *nigricaulis* seeds under *in situ* conditions may be seasonally limited by temperatures in southern Australia. Seed germination may be further restricted by salinity as freshwater pulses reaching 20 PSU are typically only observed in Port Phillip Bay following large scale rainfall events. As a result, these populations may be particularly susceptible to disturbance with only a seasonally limited capacity for recovery.

Subjects Marine Biology, Plant Science

Keywords Seagrass, Sediment, Temperature, Burial depth, Salinity, *Heterozostera tasmanica*, *Zostera nigricaulis*, Resilience

INTRODUCTION

Seagrasses are a group of marine angiosperms that are a conspicuous element of coastal environments where they stabilize sediments (*Orth et al., 2006; Ward, Michael Kemp & Boynton, 1984*), provide food and habitat for economically important recreational and commercial fisheries species (*Hemminga & Duarte, 2000*), cycle nutrients (*Den Hartog, 1970; Hemminga & Duarte, 2000; Orth et al., 2006*), promote biodiversity (*Beck et al., 2001; Orth et al., 2006; Short & Wyllie-Echeverria, 1996*) and provide long term storage of organic

Submitted 12 December 2016 Accepted 22 February 2017 Published 23 March 2017

Corresponding author Erin Cumming, erincumming1@gmail.com

Academic editor Christopher Lortie

Additional Information and Declarations can be found on page 13

DOI 10.7717/peerj.3114

Copyright 2017 Cumming et al.

Distributed under Creative Commons CC-BY 4.0

OPEN ACCESS

carbon (*Fourqurean et al., 2012*; *Orth et al., 2006*). *Zostera nigricaulis* (Kuo) Jacobs and Les. (formerly referred to as *Heterozostera tasmanica*) is a common subtidal seagrass species found in southern Australia (*Carruthers et al., 2007*; *Smith et al., 2013*). In Port Phillip Bay *Z. nigricaulis* provides a major subtidal habitat to many species, however, the species has shown significant decline over the past decade that has been attributed to environmental conditions (*Ball, Soto-Berelov & Young, 2014*; *Hirst et al., 2016*). Globally, anthropogenic and natural impacts such as coastal development and storms (*Orth et al., 2006*) have led to a 29% global loss of seagrass over the last century (*Waycott et al., 2009*). Following these declines, it has become increasingly important to understand what factors are influencing seagrass resilience to stressors as well as limiting their recovery from disturbances.

Seagrasses reproduce both sexually via flowering and asexually through rhizome extension and vegetative fragmentation (*Den Hartog, 1970; McMahon et al., 2014; Thomson et al., 2015*). The resilience of seagrass ecosystems to disturbance relies on their ability to resist environmental stressors and to recover from loss via sexual and/or asexual reproductive mechanisms (*Macreadie, York & Sherman, 2014; Orth et al., 2006; Unsworth et al., 2015*). While seagrasses have the capability to recover from disturbance through the use of rhizomes (*Frederiksen et al., 2004; Neckles et al., 2005; Rasheed, 1999*), when complete above-ground biomass has been lost, initial re-establishment is determined by sexual reproduction in relation to seed germination and seed bank density (*Greve et al., 2005; Plus, Deslous-Paoli & Dagault, 2003*). Under these conditions seed bank germination rates are important factors influencing primary natural re-establishment of sexual recruits (*Jarvis & Moore, 2010; Lee et al., 2007*).

Seed germination is a potential limiting stage in successful sexual reproduction for both terrestrial (*Harper, 1977*) and marine angiosperms (*Jarvis & Moore, 2015; Marion & Orth, 2012; Orth et al., 2000*). Germination failure has predominantly been related to the characteristics of the surrounding microenvironment which may lack the required signals to break seed dormancy, which can last up to 12 months for other *Zostera* species (*Orth et al., 2000*), and initiate germination (*Baskin & Baskin, 2014*). Therefore, understanding those environmental cues that reduce time to germination and increase the maximum number of germinated seeds is essential to determine the potential for seagrass recovery via sexual reproduction. Germination cues have recently been identified as an important knowledge gap in Australian seagrass research (*York et al., 2016*).

Seagrass germination experiments have primarily focused on two species, *Zostera marina* (*Abe, Kurashima & Maegawa, 2008; Jarvis & Moore, 2015; Marion & Orth, 2009; Moore, Orth & Nowak, 1993; Probert & Brenchley, 1999*) and *Z. muelleri (Brenchley & Probert, 1998; Conacher et al., 1994; Stafford-Bell, Chariton & Robinson, 2016*). For both *Zostera* species temperature, salinity, anoxic conditions and burial depth have been documented to significantly affect germination success (*Abe, Kurashima & Maegawa, 2008; Conacher, Poiner & O'Donohue, 1994; Jarvis & Moore, 2015; Probert & Brenchley, 1999*). In general greater germination (defined as higher maximum germination and shorter time to germination) for both species occurs at lower temperatures (5 °C–16 °C) and salinities <20 PSU and under anoxic conditions (Brenchley & Probert, 1998; Conacher et al., 1994; Moore, Orth & Nowak, 1993; Orth & Moore, 1983; Probert & Brenchley, 1999; Van Lent & Verschuure,

1995). Of the few studies that investigated the effects of sediment and burial depth on *Zostera* seed germination, seeds buried between 1 cm and 5 cm had greater maximum germination and shorter time to germination than seeds found at depths greater than 5 cm (*Wang et al., 2016*) and that time to germination was also shorter in anoxic (*Jarvis & Moore, 2010*) or fine grained sediments compared to oxygenated coarse sediments (*Wang et al., 2016*). To date, environmental cues for *Z. nigricaulis* seed germination are undefined.

Recent studies have described *Z. nigricaulis* life history and the role seeds play in meadow maintenance (*Smith et al., 2016a; Smith et al., 2016b; Thomson et al., 2015*). The aim of this study was to add to this knowledge base by determining optimal germination conditions for *Z. nigricaulis* under controlled laboratory conditions by quantifying time to germination and maximum germination of seeds across a range of temperature, salinity, burial depths and sediment conditions.

METHODS

Seed collection and storage

Zostera nigricaulis flowering shoots with mature spadices were collected in December 2012, during the period of maximum seed production (Smith et al., 2016b), from Blairgowrie (38°21'47"S, 144°47'28"E) in Port Phillip Bay, Victoria, Australia (Fig. S1). Samples were stored in flow through outdoor mesocosms ($60 \text{ L volume}, 60 \times 35 \times 30 \text{ cm}$) at the Victorian Marine Science Consortium (VMSC) in Queenscliff, Victoria under ambient conditions until seeds dehisced from the reproductive shoots. Seagrass samples were then sorted by hand and sieved (710 μ m mesh) to separate mature seeds from remaining vegetative material (Marion & Orth, 2010). After separation, seeds were stored indoors in 1 L tanks with flow through seawater (~21 °C) and light aeration (Jarvis & Moore, 2015; Jarvis & Moore, 2010). All seeds used in experiments were scarified using a scalpel under a dissecting microscope to create a slight opening in the seed coat to stimulate germination (*Karrfalt, 2008*). Any seeds that appeared damaged after scarification (e.g., embryo visible through the seed coat) or to have developed fungal growth were discarded. Previous research has shown few seeds with intact seed coats germinate (Conacher et al., 1994), therefore we chose to scarify seeds to promote seed germination and allow the effect of different environmental conditions on germination to be tested.

Sediment collection and characterization

Three replicate sediment cores (2 cm width \times 3 cm height) were collected from three sites (n = 9) which represented a range of observed different sediment types within *Z. nigricaulis* meadows (Blairgowrie) Avalon (38°05′08″S, 144°25′43″E) and Williamstown (37°52′14″S, 144°54′32″E) and stored in a cool room (4 °C) until processing. All sediment samples were analysed for percent organic matter and sediment grain size was quantified using standard methods (*Erftemeijer & Koch, 2001*). Organic matter was measured by drying sediment sample cores in a drying oven at 60 °C for 24 h followed by 5 h in a blast furnace (500 °C), with percentage of organic matter lost on ignition recorded after being weighed. To quantify grain size, sediment samples were exposed to hydrogen peroxide (30%) for 24 h to eliminate organic matter, weighed and sieved (>2 mm, 1–2 mm, 500 µm–1 mm,

250–500 μm, 125–250 μm, 62–125 μm, <62 μm fractions) and the weight of each amount of sediment left in each sieve was recorded (*Erftemeijer & Koch*, 2001).

Temperature and salinity experiment

Maximum germination and time to germination of *Z. nigricaulis* seeds were assessed across a 3-way fully orthogonal design with treatments of temperature (3 levels: 13 °C, 17 °C, 22 °C), salinity (3 levels; 25 PSU, 30 PSU, 35 PSU) and low salinity pulse (2 levels: 24 h pulse in 20 PSU seawater and no pulse). Temperature and salinity concentrations were chosen to reflect the natural variation found in Port Phillip Bay, while the low salinity pulse was chosen to represent stressful environmental conditions (*Lee et al., 2012*; *Walker, 1999*). Fifty seeds were randomly allocated to one of four replicate petri dishes for each of the 18 treatments containing damp filter paper and 3–5 ml of saline solution and placed into a temperature control room with 12 h light cycles. Salinity of treatments was monitored daily and saline solution was added when necessary (every 2nd or 3rd day). Seeds were scored as successfully germinated when the cotyledon was extended 0.5 mm or more from the seed (*Conacher et al., 1994*; *Jarvis & Moore, 2010*). The number of germinated seeds and the salinity of each treatment was recorded weekly at the beginning of the experiment and then fortnightly until completion of the experiment 107 days later.

Burial depth and sediment composition experiment

Following the completion of the salinity and temperature experiment, maximum germination of *Z. nigricaulis* seeds in sediment was assessed across a 2-way fully orthogonal design with treatments of burial depth (3 levels: 0 cm, 1 cm, 2 cm) and varying sediment composition based on grain size distribution (3 levels: fine (>25 % fine sediment), medium (10% fine sediment), and coarse (<5% fine sediment)). Sediment from each site was then divided into twelve 11 × 6 cm plastic experimental cores and 25 seeds were buried at the allocated depths. Based on the results of the previous experiment, all seeds were exposed to a fresh water pulse (20% PSU for 48 h) before burial to maximize germination. All cores were then randomly placed into 2 tanks (100 × 40 cm) with flow-through seawater (~35 PSU). Tanks were held in a temperature control room (13 °C) with a 12 h light cycle for 7 weeks. Germination (cotyloid growth of 0.5 mm) was recorded every two weeks (*Conacher et al.*, *1994*; *Jarvis & Moore*, *2010*) until the completion of the experiment after 50 days.

Statistical analysis

Prior to the beginning of the experiments, a set of analytical models were developed to describe the relationship between maximum germination and mean time to germination (MTG) for each experiment. To determine the best fitting model, the Akaike's Information Criterion corrected for small sample sizes (AICc) was calculated using loglikelihood ratios derived from all regression analyses (*Burnham & Anderson, 2002*). AICc differences between all models were then calculated and the models were ranked (*Barton, 2013*; *R Core Team, 2014*). The best-fitting model was considered to be the simplest model that fell within two of the lowest AICc (*Burnham & Anderson, 2002*; Tables S1–S5). Overall effects of categorical variables on MTG and maximum germination were calculated with Wald Chi square tests using the 'Imtest' package (*Zeileis & Hothorn, 2002*).

Based on the large numbers of zeros found within the treatments (\sim 43%) and overall low germination response in both experiments, the effects of experimental treatments on maximum germination of *Z. nigricaulis* seeds were analysed using a zero/one inflated beta binomial (ZOIB) regression with the 'gamlss' package (*Rigby & Stasinopoulos, 2005*) in the statistical program R (*R Core Team, 2014*). ZOIB regression models can be used to model response variables that are bound between or equal to 0 or 1 and contain a non-negligible number of zeros and or ones (*Ospina & Ferrari, 2010*).

Due to the potential for a large amount of right-censored data characteristic of germination experiments (McNair, Sunkara & Frobish, 2012; Scott, Jones & Williams, 1984), Cox models were selected to quantify treatment effects (pulse and non-pulsed seeds, temperature and salinity, burial depth and sediment source) on mean time to germination (MTG). Seed data were censored if seed germination did not occur and non-germinated seeds were flagged prior to analysis. As time-event analyses are based on the distribution of germination times of individual seeds rather than on cumulative germination curves, each seed was analysed independently using the survival package (Therneau & Grambsch, 2000). Germination data were first graphically explored for violations of the proportional hazards function by plotting separate non-parametric Kaplan-Meier survivorship functions for the different factors (McNair, Sunkara & Frobish, 2012), KM surv (Klein & Moeschberger, 2005). Potential multicollinearity of the covariates were tested by calculating variance inflation factors (Heiberger & Holland, 2015) for all treatment factors prior to analysis. The effects of treatment on time to germination were then calculated using the Cox model (survival, Therneau & Grambsch, 2000). Seed germination was only recorded at the end of the sediment experiment and therefore MTG was not calculated.

Sediment grain size and percentage organic matter data were calculated as proportions of total sample weight and independently analysed using SYSTAT 12 with a one-way analysis of variance to compare between treatments (*Quinn & Keough, 2002*). Prior to analysis all sediment data were transformed when necessary to meet the assumptions of normality and homogeneity of variance (*Quinn & Keough, 2002*). All post hoc analyses of the data were performed with Tukey's test.

RESULTS

Salinity and temperature experiment

Overall, maximum germination across all treatments was low (<20%; Table 1), however, seeds were more likely to germinate when exposed to a low salinity pulse (p < 0.001; Table 2). There was a significant interaction between pulse and salinity treatments (Wald chi square test, p = 0.047) with germination decreasing with increasing salinity in the non-pulsed treatment and no effect of salinity on germination in the seeds exposed to a low salinity pulse (Fig. 1 and Table 2). The inclusion of temperature in the model did not improve model fit and therefore was removed from the maximum germination analysis (Table S1).

Mean time to germination ranged from 37 ± 1 to 76 ± 1 days across all treatments. There was a significant interaction on mean time to germination between salinity, temperature and

Pulse	Temp (°C)	Salinity	MTG (days)	Max % G
Yes	25	13	37 ± 1	10 ± 2
		17	56 ± 1	9 ± 3
		22	50 ± 1	5 ± 1
	30	13	51 ± 1	12 ± 6
		17	54 ± 2	16 ± 12
		22	44 ± 1	10 ± 7
	35	13	54 ± 1	14 ± 4
		17	48 ± 1	9 ± 3
		22	42 ± 1	5 ± 1
No	25	13	42 ± 1	10 ± 1
		17	55 ± 2	4 ± 2
		22	47 ± 1	5 ± 2
	30	13	76 ± 3	3 ± 1
		17	71 ± 2	6 ± 3
		22	48 ± 1	4 ± 2
	35	13	43 ± 2	1 ± 1
		17	71 ± 2	4 ± 2
		22	73 ± 3	3 ± 1

 Table 1
 Maximum germination (Max % G) and mean time to germination (MTG) across all treatments in the salinity and temperature experiment (±SE).

 Table 2
 Zero/one inflated beta regression results for maximum germination of Z. nigricaulis seeds across pulse, salinity and temperature treatments.

Parameter	Coef.	SE	<i>t</i> value	<i>p</i> -value
Intercept	-3.542	0.289	-12.256	< 0.001*
Pulse	1.300	0.343	3.793	< 0.001*
Salinity 25 PSU	-	-	-	-
Salinity 30 PSU	0.633	0.346	1.832	0.072
Salinity 35 PSU	0.881	0.354	2.488	0.016*
Pulse:Sal (25 PSU)	-	-	-	_
Pulse:Sal (30 PSU)	-0.284	0.467	0.608	0.545
Pulse:Sal (35 PSU)	-0.930	0.459	-2.026	0.047^{*}

Notes.

*indicates a significant value.

low salinity pulse treatments (p = 0.023, Fig. 2 and Table 3). Mean time to germination was on average 10 days earlier in the pulsed treatment (48 ± 6 days) compared to the non-pulsed treatment (58 ± 14 days) and in low salinity (48 ± 17 days) compared to medium (57 ± 13 days) and high salinity treatments (55 ± 14 days) (Table 1). Mean time to germination was shorter in the 13 °C (51 ± 14 days) and 22 °C (51 ± 11 days) treatments compared to 17 °C (59 ± 10 days) (Table 1). The shortest MTG (37 ± 1 day) occurred with seeds exposed to a low salinity pulse and exposed to salinities of 25 PSU and temperatures of 13 °C (Table 1).







Figure 2 Maximum germination of *Z. nigricaulis* seeds for the salinity and temperature experiment. White bars indicate the 25 PSU treatment, light grey is 30 PSU and the dark grey represent the 35 PSU treatment.

Parameter	Coef.	Exp	SE	Z score	<i>p</i> -value	
Pulse	-12.540	0.000	5.033	-2.492	0.013*	
Salinity	-0.402	0.669	0.150	-2.691	0.007^{*}	
Temperature	-0.518	0.596	0.240	-2.162	0.031*	
Pulse: Salinity	0.493	1.637	0.174	2.840	0.005*	
Pulse: Temp	0.596	1.816	0.287	2.080	0.038	
Salinity: Temp	0.018	1.018	0.008	2.119	0.034*	
Pulse: Salinity: Temp	-0.022	0.978	0.010	-2.277	0.023*	

 Table 3
 Results of survival analysis looking at differences in the main effects for pulse, salinity and temperature on mean time to germination of Z. nigricaulis seeds.

Notes.

*indicates a significant value.

Sediment type and burial depth experiment

Organic matter was significantly higher at Avalon than Williamstown and Blairgowrie $(F_{2,6} = 8.51, p = 0.018, \text{Fig. 3A})$. Of the seven sediment grain sizes measured only two showed any significant difference across sites. The proportion of medium grain sand (250–500 µm) was highest at Blairgowrie, followed by Avalon and Williamstown $(F_{2,6} = 43.9, p < 0.001, \text{Fig. 3B})$. Very fine grain sand (65–125 µm) was highest at Williamstown followed by Avalon and Blairgowrie which were not significantly different (p = 0.054). Therefore, based on organic matter content and grain size distribution, Williamstown sediment was used for the 'fine' sediment treatment, Avalon for the 'medium' and Blairgowrie for 'coarse' treatments.

When placed in sediment, maximum germination fell below 10% regardless of burial depth or sediment type (Table 4). Germination counts were not made frequently enough to calculate mean time to germination; however, maximum germination was significantly affected by the interaction between sediment type and burial depth (Wald Chi square test, p = 0.009). While there was no significant difference in germination between seeds buried at 0 and 1 cm (p = 0.660) or between seeds at 1 and 2 cm (p = 0.721), seeds placed on the sediment surface (0 cm) had a greater maximum germination than seeds buried at 2 cm (p = 0.048; Fig. 3; Table 5). Overall, seeds in Williamstown and Avalon sediment had greater germination than seeds placed in Blairgowrie sediment except at 2 cm depths (Fig. 4; Table 5).

DISCUSSION

Surprisingly, overall germination of *Z. nigricaulis* seeds across all experiments was low (<20%) compared to other *Zostera* species. Germination rates of *Z. muelleri* (formerly *Z. capricorni*), a co-occurring intertidal species, range between 20%–60% under similar conditions (*Brenchley & Probert, 1998; Conacher et al., 1994*) while temperate northern hemisphere species *Z. marina* and *Z. noltii* range from 5%–100% and <10%–80% respectively (*Hootsmans, Vermaat & Van Vierssen, 1987; Van Lent & Verschuure, 1995*). However, high germination rates in many of these studies were at very low salinity levels (<20 PSU) that are rarely encountered in the field (*Lee et al., 2012; Probert & Brenchley, 1999*). When



Figure 3 Experimental sediment characterization. (A) Mean (\pm SE) % organic matter content at Avalon, Williamstown and Blairgowrie and (B) mean (\pm SE) % of medium (250–500 µm) and fine (62–125 µm) grain sediment at Avalon, Williamstown and Blairgowrie.

Sediment type	Burial depth (cm)	Max % G
Fine	0	6 ± 4
	1	8 ± 3
	2	1 ± 1
Medium	0	5 ± 4
	1	4 ± 4
	2	3 ± 1
Coarse	0	3 ± 2
	1	3 ± 1
	2	2 ± 1

Table 4Maximum Z. nigricaulis germination (% G) across all treatments in the sediment type andburial depth experiment.Values are given as mean \pm SE.



Figure 4 Maximum germination of *Z. nigricaulis* seeds (mean \pm SE) for sediment type and burial **depth experiment.** The white bars represent fine, light grey represent medium and the dark grey bars represent coarse sediment grain size treatments.

considering salinities of 25–40 PSU, which overlap with this experiment and are more reflective of natural conditions observed in Port Phillip Bay, *Z. muelleri* germination was <20% in all treatments except when temperatures were <16 °C (*Brenchley & Probert, 1998*; *Conacher et al., 1994*). While germination was also greatest at low temperatures in this study (25 PSU at 13 °C), the overall low germination response indicates that additional

Parameter	Coef.	SE	t value	<i>p</i> -value
Intercept	-2.013	0.214	-9.391	< 0.001*
Fine	_	-	-	-
Medium	-0.232	0.2287	-0.810	0.426
Coarse	-0.719	0.355	-2.028	0.054
Depth 0 cm	-0.126	0.282	-0.446	0.660
Depth 1 cm	-0.126	0.282	-0.446	0.660
Depth 2 cm	-1.055	0.505	-2.088	0.048^{*}
Fine: depth (1)	_	-	-	-
Medium: depth (1)	0.006	0.419	0.014	0.989
Coarse: depth (1)	-0.210	0.478	-0.439	0.665
Fine: depth (2)	-	-	-	-
Medium: depth (2)	1.667	0.602	2.769	0.011*
Coarse: depth (2)	0.719	0.663	1.085	0.289

Table 5Zero/one inflated beta regression results for maximum germination of Z. nigricaulis seeds forsediment type and burial depth treatments.

Notes.

*indicates a significant value.

germination cues (e.g., dissolved oxygen, light, variations in sediment microbial communities) both individually or in combination may be missing and require further investigation.

Salinity and temperature are key germination cues for many Zostera species (Conacher et al., 1994; Hootsmans, Vermaat & Van Vierssen, 1987; Kaldy et al., 2015; Orth & Moore, 1983; Stafford-Bell, Chariton & Robinson, 2016; Van Lent & Verschuure, 1995). Seeds of Z. nigricaulis generally had greater and quicker germination in lower salinities and temperatures. Although consistent with other germination studies recorded for Z. muelleri, Z. marina and Z. noltii (Brenchley & Probert, 1998; Conacher et al., 1994; Probert & Brenchley, 1999; Van Lent & Verschuure, 1995), high germination at low salinities seems an unlikely germination cue for Z. nigricaulis. Zostera nigricaulis is found in large bays and protected coastal habitats but is absent from estuaries suggesting it has little tolerance of low salinities. Lower salinities promoted higher and faster germination rates and in most cases a low salinity pulse increased germination. In contrast, Z. muelleri and Z. marina are often found in estuaries and therefore greater germination at low salinities may have important ecological implications. Salinities in Port Phillip Bay rarely reach the level of 20-30 NTU that was used for a 24 h pulse and as a low salinity treatment and there are few freshwater inputs to reduce salinities under flood conditions (Lee et al., 2012; Walker, 1999). Flood conditions that lower salinities and cause disturbance creating space for seeds to grow into could explain high germination in low salinity; however, floods are generally associated with high sediment and turbidity levels that may bury seeds and restrict light for seedling growth.

Variation in levels of *Z. nigricaulis* seed germination at different temperatures and salinities may have important ecological implications. Temperature and salinity values used in this study were chosen to reflect the range within Port Phillip Bay. Peak flowering in *Z. nigricaulis* occurs during October and November each year (*Smith et al., 2016b*), which coincides with mean seawater temperatures between 13 °C and 17 °C. Once germinated, seedling growth may be rapid at this time of the year as temperatures and daylight increase moving into the Austral summer. Germinating at the onset of optimal growing conditions increases the length of the peak growing period for seedlings increasing their chance of survival. This reflects seed germinations in other species that coincides with high seed banks and proceeds the growing season (*Garwood, 1983*).

Changes in global climate conditions will impact plant species reproduction and resilience strategies. In Port Phillip Bay, water temperature and salinity are expected to increase over the next 15 years with salinity expected to increase by as much as 4 NTU in the Geelong Arm of Port Phillip Bay (*Lee et al., 2012*). Already the Geelong Arm has sustained considerable seagrass loss (*Ball, Soto-Berelov & Young, 2014*) and further increases in salinity will restrict the ability of seagrass to recover from seeds as germination decreases. Seeds are often an important recovery mechanism in seagrass ecosystems (*Alexandre, Santos & Serrao, 2005; Hammerstrom et al., 2006; Jarvis & Moore, 2010*) but changing conditions in the future will have major implications for the ability of seagrass to recover and reduce the resilience of seagrass habitats. To fully understand the impacts of environmental changes in near shore coastal environments, and to gain a better understanding of potential resilience of this species to increased disturbance, additional research is required to better understand seed ecology.

Sediment conditions

In addition to salinity and temperature, germination of Z. nigricaulis seeds is affected by both burial depth and sediment type, consistent with previous studies in other Zostera species (Jarvis & Moore, 2015; Wang et al., 2016). Germination in fine and medium sediment was greater when seeds were at the sediment surface than when they were buried at 2 cm. Thus it is clear from this and previous studies that burial depth can affect Zostera germination (Granger, Traber & Nixon, 2000; Jarvis & Moore, 2015) and suggests that either germination cues interact with burial depth to affect germination rates, or that seeds germinate but do not have the energy stores required for the cotyledon to reach the surface (Jarvis & Moore, 2015; Wang et al., 2016). Seagrass beds accumulate sediment and organic matter (Bos et al., 2007; Fonseca & Fisher, 1986), burying seeds and consequently reducing germination rates. Therefore, although there may be a significant seed bank, as sediment is deposited over time, the likelihood of germination is lower and potentially decreases the ability of patches to recover from disturbances. Similarly, large scale along shore sediment movement can play a role in seagrass distributions and seagrass loss can occur through burial (Ball, Soto-Berelov & Young, 2014). These results suggest that in such a situation recovery is unlikely to occur from the seed bank.

Germination in coarse sediment was lower than in other grain sizes and showed no difference across depths. Variations in germination in different sediment types can be attributed to a variety of factors such as organic matter, grain size and anoxia (*Jarvis & Moore, 2015; Tanner & Parham, 2010; Wang et al., 2016*). High organic matter in sediment can lead to anoxic conditions, which can have higher germination rates than aerobic conditions (*Brenchley & Probert, 1998; Probert & Brenchley, 1999*). Organic matter in the coarse sediment was lower than at the other sites and therefore may be playing an important

role in determining germinations rates. Likewise, sediment grain size is thought to affect sediment nutrient levels and can affect time to germination (*Jarvis & Moore, 2015; Wang et al., 2016*). Finer sediment in seagrass beds reduces pore water loss, increasing nutrient levels while in coarse sediment pore water, and nutrients are easily lost (*Koch, 2001*). Variations in nutrient levels in the different sediment treatments may explain differences in *Z. nigricaulis* seed germination but the role sediment nutrients levels play in seagrass seed germination is unknown (*Jarvis & Moore, 2015*). Further research into the impacts of varying sediment conditions are needed to explain why germination rates vary across sediment conditions.

Like many plant species seagrasses invest large amounts of energy into the production of seeds often producing vast quantities that enter the seed bank (e.g., *Smith et al., 2016b*). Seed banks are important for recovery and maintenance in many ecosystems that are susceptible to local habitat loss. Low germination rates may impact the resilience of *Z. nigricaulis* to disturbance and future environmental change. Inability to recover can have implications for coastal ecosystems given the many ecosystem services and therefore more research to better understand seagrass seed ecology.

ACKNOWLEDGEMENTS

Rod Watson assisted with field collection and laboratory experiments. All work was done at the Victorian Marine Science Consortium (VMSC).

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

The authors received no funding for this work.

Competing Interests

The authors declare there are no competing interests.

Author Contributions

- Erin Cumming conceived and designed the experiments, performed the experiments, analyzed the data, contributed reagents/materials/analysis tools, wrote the paper, prepared figures and/or tables.
- Jessie C. Jarvis analyzed the data, wrote the paper, prepared figures and/or tables, reviewed drafts of the paper.
- Craig D.H. Sherman conceived and designed the experiments, contributed reagents/materials/analysis tools, wrote the paper, reviewed drafts of the paper.
- Paul H. York conceived and designed the experiments, performed the experiments, wrote the paper, reviewed drafts of the paper.
- Timothy M. Smith conceived and designed the experiments, performed the experiments, analyzed the data, contributed reagents/materials/analysis tools, wrote the paper, prepared figures and/or tables, reviewed drafts of the paper.

Data Availability

The following information was supplied regarding data availability:

Cumming, Erin and Smith, Timothy 2016, Zostera nigricaulis seed germination experiment [data collection], doi: 10.4225/16/5848b8f8d273b.

Supplemental Information

Supplemental information for this article can be found online at http://dx.doi.org/10.7717/ peerj.3114#supplemental-information.

REFERENCES

- Abe M, Kurashima A, Maegawa M. 2008. Temperature requirments for seed germination and growth of *Zostera marina* from central Japan. *Fisheries Science* 74:589–593 DOI 10.1111/j.1444-2906.2008.01562.x.
- Alexandre A, Santos R, Serrao E. 2005. Effects of clam harvesting on sexual reproduction of the seagrass *Zostera noltii*. *Marine Ecology Progress Series* 298:115–122 DOI 10.3354/meps298115.
- Ball D, Soto-Berelov M, Young P. 2014. Historical seagrass mapping in Port Phillip Bay, Australia. *Journal of Coastal Conservation* 18:257–272 DOI 10.1007/s11852-014-0314-3.
- Barton K. 2013. MuMIn: multi-model inference. *Available at https://CRAN.R-project. org/package=MuMIn*.
- **Baskin C, Baskin J. 2014.** *Seeds: ecology, biogeography, and evolution of dormancy and germination.* 2nd edition. San Diego: Academic Press.
- Beck MW, Heck Jr KL, Able KW, Childers DL, Eggleston DB, Gillanders BM, Halpern
 B, Hays CG, Hoshino K, Minello TJ, Orth RJ, Sheridan PF, Weinstein MP. 2001.
 The identification, conservation, and management of estuarine and marine nurseries for fish and invertebrates. *BioScience* 51:633–641
 DOI 10.1641/0006-3568(2001)051[0633:TICAMO]2.0.CO;2.
- Bos AR, Bouma TJ, De Kort GLJ, Van Katwijk MM. 2007. Ecosystem engineering by annual intertidal seagrass beds: sediment accretion and modification. *Estuarine, Coastal and Shelf Science* 74:344–348 DOI 10.1016/j.ecss.2007.04.006.
- Brenchley JL, Probert RJ. 1998. Seed germination responses to some environmental factors in the seagrass *Zostera capricorni* from eastern Australia. *Aquatic Botany* 62:177–188 DOI 10.1016/S0304-3770(98)00089-8.
- **Burnham KP, Anderson DR. 2002.** Information and likelihood theory: a basis for model selection and inference. In: *Model selection and multimodel inference: a practical information-theoretic approach.* New York: Springer-Verlag, 49–97.
- Carruthers TJB, Dennison WC, Kendrick GA, Waycott M, Walker DI, Cambridge ML. 2007. Seagrasses of south-west Australia: a conceptual synthesis of the world's most diverse and extensive seagrass meadows. *Journal of Experimental Marine Biology and Ecology* 350:21–45 DOI 10.1016/j.jembe.2007.05.036.

- **Conacher CA, Poiner IR, Butler J, Pun S, Tree DJ. 1994.** Germination, storage and viability testing of seeds of *Zostera capricorni* Aschers. from a tropical bay in Australia. *Aquatic Botany* **49**:47–58 DOI 10.1016/0304-3770(94)90005-1.
- **Conacher CA, Poiner IR, O'Donohue M. 1994.** Morphology, flowering and seed production of *Zostera capricorni* Aschers. in subtropical Australia. *Aquatic Botany* **49**:33–46 DOI 10.1016/0304-3770(94)90004-3.

Den Hartog C. 1970. Seagrasses of the world. Amsterdam: New Holland.

- **Erftemeijer PLA, Koch EW. 2001.** Sediment geology methods for seagrass habitat. In: Short FT, Coles RG, eds. *Global seagrass research methods*. Amsterdam: Elsevier Science B.V., 345–367.
- **Fonseca MS, Fisher JS. 1986.** A comparison of canopy friction and sediment movement between four species of seagrass with reference to their ecology and restoration. *Marine Ecology Progress Series* **29**:15–22 DOI 10.3354/meps029015.
- Fourqurean JW, Duarte CM, Kennedy H, Marbà N, Holmer M, Mateo MA, Apostolaki ET, Kendrick GA, Krause-Jensen D, McGlathery KJ. 2012. Seagrass ecosystems as a globally significant carbon stock. *Nature Geoscience* 5:505–509 DOI 10.1038/ngeo1477.
- **Frederiksen M, Krause-Jensen D, Holmer M, Laursen JS. 2004.** Long-term changes in area distribution of eelgrass (*Zostera marina*) in Danish coastal waters. *Aquatic Botany* **78**:167–181 DOI 10.1016/j.aquabot.2003.10.002.
- Garwood NC. 1983. Seed germination in a seasonal tropical forest in Panama: a community study. *Ecological Monographs* 53:159–181 DOI 10.2307/1942493.
- **Granger S, Traber M, Nixon S. 2000.** The influence of planting depth and density on germination and development of *Zostera marina* L. seeds. *Biologia Marina Mediterranea* 7:55–58.
- Greve TM, Krause-Jensen D, Rasmussen MB, Christensen PB. 2005. Means of rapid eelgrass (*Zostera marina* L.) recolonisation in former dieback areas. *Aquatic Botany* 82:143–156 DOI 10.1016/j.aquabot.2005.03.004.
- Hammerstrom KK, Kenworthy WJ, Fonseca MS, Whitfield PE. 2006. Seed bank, biomass, and productivity of *Halophila decipiens*, a deep water seagrass on the west Florida continental shelf. *Aquatic Botany* **84**:110–120 DOI 10.1016/j.aquabot.2005.08.002.
- Harper JL. 1977. Population biology of plants. London: Academic Press.
- **Heiberger RM, Holland B. 2015.** *Statistical analysis and data display.* New York: Springer, 1104 pp.
- Hemminga M, Duarte CM. 2000. *Seagrass ecology*. Cambridge: Cambridge University Press.
- Hirst AJ, Longmore AR, Ball D, Cook PLM, Jenkins GP. 2016. Linking nitrogen sources utilised by seagrass in a temperate marine embayment to patterns of seagrass change during drought. *Marine Ecology Progress Series* 549:79–88 DOI 10.3354/meps11708.
- Hootsmans MJM, Vermaat JE, Van Vierssen W. 1987. Seed-bank development, germination and early seedling survival of two seagrass species from The Netherlands:

Zostera marina L. and *Zostera noltii* hornem. *Aquatic Botany* **28**:275–285 DOI 10.1016/0304-3770(87)90005-2.

- Jarvis JC, Moore KA. 2010. The role of seedlings and seed bank viability in the recovery of Chesapeake Bay, USA, *Zostera marina* populations following a large-scale decline. *Hydrobiologia* 649:55–68 DOI 10.1007/s10750-010-0258-z.
- Jarvis J, Moore K. 2015. Effects of seed source, sediment type, and burial depth on mixed-annual and perennial *Zostera marina* L. seed germination and seedling establishment. *Estuaries and Coasts* 38:964–978 DOI 10.1007/s12237-014-9869-3.
- Kaldy JE, Shafer DJ, Ailstock MS, Magoun AD. 2015. Effects of temperature, salinity and seed age on induction of *Zostera japonica* germination in North America, USA. *Aquatic Botany* 126:73–79 DOI 10.1016/j.aquabot.2015.06.006.
- Karrfalt RP. 2008. Seed testing. In: Bonner FT, Karrfalt RP, eds. *The woody plant seed manual. USDA forest service, Agriculture Handbook number 727.* Washington, D.C.: USDA, 97–116.
- Klein JP, Moeschberger ML. 2005. Survival analysis: techniques for censored and truncated data. New York: Springer.
- **Koch EW. 2001.** Beyond light: physical, geological, and geochemical parameters as possible submersed aquatic vegetation habitat requirements. *Estuaries* **24**:1–17 DOI 10.2307/1352808.
- Lee R, Black K, Bosserel C, Greer D. 2012. Present and future prolonged drought impacts on a large temperate embayment: Port Phillip Bay, Australia. *Ocean Dynamics* 62:907–922 DOI 10.1007/s10236-012-0538-4.
- Lee K-S, Park J-I, Kim YK, Park SR, Kim J-H. 2007. Recolonization of *Zostera marina* following destruction caused by a red tide algal bloom: the role of new shoot recruitment from seed banks. *Marine Ecology Progress Series* 342:105–115 DOI 10.3354/meps342105.
- Macreadie PI, York PH, Sherman CDH. 2014. Resilience of *Zostera muelleri* seagrass to small-scale disturbances: the relative importance of asexual versus sexual recovery. *Ecology and Evolution* 4:450–461 DOI 10.1002/ece3.933.
- Marion SR, Orth RJ. 2009. Factors influencing seedling establishment rates in *Zostera marina* and their implications for seagrass restoration. *Restoration Ecology* 18:549–559 DOI 10.1111/j.1526-100X.2010.00695.x.
- Marion SR, Orth RJ. 2010. Innovative techniques for large-scale seagrass restoration using *Zostera marina* (eelgrass) seeds. *Restoration Ecology* 18:514–526 DOI 10.1111/j.1526-100X.2010.00692.x.
- Marion SR, Orth RJ. 2012. Seedling establishment in eelgrass: seed burial effects on winter losses of developing seedlings. *Marine Ecology Progress Series* 448:197–207 DOI 10.3354/meps09612.
- McMahon K, Van Dijk K-J, Ruiz-Montoya L, Kendrick GA, Krauss SL, Waycott M, Verduin J, Lowe R, Statton J, Brown E, Duarte C. 2014. The movement ecology of seagrasses. *Proceedings of the Royal Society B: Biological Sciences* 281 DOI 10.1098/rspb.2014.0878.

- McNair JN, Sunkara A, Frobish D. 2012. How to analyse seed germination data using statistical time-to-event analysis: non-parametric and semi-parametric methods. *Seed Science Research* 22:77–95 DOI 10.1017/S0960258511000547.
- Moore KA, Orth RJ, Nowak JF. 1993. Environmental regulation of seed germination in *Zostera marina* L. (eelgrass) in Chesapeake Bay: effects of light, oxygen and sediment burial. *Aquatic Botany* **45**:79–91 DOI 10.1016/0304-3770(93)90054-Z.
- Neckles HA, Short FT, Barker S, Kopp BS. 2005. Disturbance of eelgrass *Zostera marina* by commercial mussel *Mytilus edulis* harvesting in Maine: dragging impacts and habitat recovery. *Marine Ecology Progress Series* 285:57–73 DOI 10.3354/meps285057.
- Orth RJ, Carruthers TJB, Dennison WC, Duarte CM, Fourqurean JW, Heck KL, Hughes AR, Kendrick GA, Kenworthy WJ, Olyarnik S, Short FT, Waycott M, Williams SL. 2006. A global crisis for seagrass ecosystems. *BioScience* 56:987–996 DOI 10.1641/0006-3568(2006)56[987:AGCFSE]2.0.CO;2.
- Orth RJ, Harwell MC, Bailey EM, Bartholomew A, Jawad JT, Lombana AV, Moore KA, Rhode JM, Woods HE. 2000. A review of issues in seagrass seed dormancy and germination: implications for conservation and restoration. *Marine Ecology Progress Series* 200:277–288 DOI 10.3354/meps200277.
- Orth RJ, Moore KA. 1983. Seed germination and seedling growth of *Zostera marina* L. (eelgrass) in the chesapeake bay. *Aquatic Botany* 15:117–131 DOI 10.1016/0304-3770(83)90023-2.
- Ospina R, Ferrari SL. 2010. Inflated beta distributions. *Statistical Papers* 51:111–126 DOI 10.1007/s00362-008-0125-4.
- Plus M, Deslous-Paoli JM, Dagault F. 2003. Seagrass (*Zostera marina* L.) bed recolonisation after anoxia-induced full mortality. *Aquatic Botany* 77:121–134 DOI 10.1016/S0304-3770(03)00089-5.
- **Probert RJ, Brenchley JL. 1999.** The effect of environmental factors on field and laboratory germination in a population of *Zostera marina* L. from southern England. *Seed Science Research* **9**:331–339.
- **Quinn GP, Keough MJ. 2002.** *Experimental design and data analysis for biologists.* Cambridge: Cambridge University Press.
- **R Core Team. 2014.** R: a language and environment for statistical computing. Vienna: R Foundation for Statistical Computing.
- Rasheed MA. 1999. Recovery of experimentally created gaps within a tropical *Zostera capricorni* (Aschers.) seagrass meadow, Queensland Australia. *Journal of Experimental Marine Biology and Ecology* 235:183–200 DOI 10.1016/S0022-0981(98)00158-0.
- **Rigby RA, Stasinopoulos DM. 2005.** Generalized additive models for location, scale and shape. *Journal of the Royal Statistical Society: Series C* **54**:507–554 DOI 10.1111/j.1467-9876.2005.00510.x.
- Scott SJ, Jones RA, Williams WA. 1984. Review of data analysis methods for seed germination. Crop Science 24:1192–1199

DOI 10.2135/cropsci1984.0011183X002400060043x.

Short FT, Wyllie-Echeverria S. 1996. Natural and human-induced disturbance of seagrasses. *Environmental Conservation* 23:17–27 DOI 10.1017/S0376892900038212.

- Smith TM, York PH, Macreadie PI, Keough MJ, Ross DJ, Sherman CDH. 2016a. Recovery pathways from small-scale disturbance in a temperate Australian seagrass. *Marine Ecology Progress Series* 542:97–108 DOI 10.3354/meps11531.
- Smith TM, York PH, Macreadie PI, Keough MJ, Ross DJ, Sherman CDH. 2016b. Spatial variation in reproductive effort of a southern Australian seagrass. *Marine Environmental Research* 120:214–224 DOI 10.1016/j.marenvres.2016.08.010.
- Smith TM, York PH, Stanley AM, Macreadie PI, Keough MJ, Ross DJ, Sherman CH. 2013. Microsatellite primer development for the seagrass *Zostera nigricaulis* (Zoster-aceae). *Conservation Genetics Resources* 5:607–610 DOI 10.1007/s12686-013-9862-3.
- Stafford-Bell RE, Chariton AA, Robinson RW. 2016. Germination and early-stage development in the seagrass, *Zostera muelleri* Irmisch ex Asch. in response to multiple stressors. *Aquatic Botany* 128:18–25 DOI 10.1016/j.aquabot.2015.09.004.
- Tanner CE, Parham T. 2010. Growing *Zostera marina* (eelgrass) from seeds in landbased culture systems for use in restoration projects. *Restoration Ecology* 18:527–537 DOI 10.1111/j.1526-100X.2010.00693.x.
- **Therneau TM, Grambsch PM. 2000.** *Modeling survival data: extending the Cox model.* Berlin: Springer.
- Thomson ACG, York PH, Smith TM, Sherman CDH, Booth DJ, Keough MJ, Ross DJ, Macreadie PI. 2015. Seagrass viviparous propagules as a potential long-distance dispersal mechanism. *Estuaries and Coasts* 38:927–940 DOI 10.1007/s12237-014-9850-1.
- Unsworth RKF, Collier CJ, Waycott M, McKenzie LJ, Cullen-Unsworth LC. 2015. A framework for the resilience of seagrass ecosystems. *Marine Pollution Bulletin* 100:34–46 DOI 10.1016/j.marpolbul.2015.08.016.
- Van Lent F, Verschuure JM. 1995. Comparative study on populations of *Zostera marina* L. (eelgrass): experiemntal germination and growth. *Journal of Experimental Marine Biology and Ecology* 185:77–91 DOI 10.1016/0022-0981(94)00132-W.
- Walker SJ. 1999. Coupled hydrodynamic and transport models of Port Phillip Bay, a semi-enclosed bay in south-eastern Australia. *Marine and Freshwater Research* 50:469–481 DOI 10.1071/MF98071.
- Wang M, Wang Y, Guo X, Sha J, Zhang H, Tang X, Zhou B. 2016. Reproductive properties of *Zostera marina* and effects of sediment type and burial depth on seed germination and seedling establishment. *Aquatic Botany* **134**:68–74 DOI 10.1016/j.aquabot.2016.07.003.
- Ward LG, Michael Kemp W, Boynton WR. 1984. The influence of waves and seagrass communities on suspended particulates in an estuarine embayment. *Marine Geology* 59:85–103 DOI 10.1016/0025-3227(84)90089-6.
- Waycott M, Duarte CM, Carruthers TJB, Orth RJ, Dennison WC, Olyarnik S, Calladine A, Fourqurean JW, Heck KL, Hughes AR, Kendrick GA, Kenworthy WJ, Short FT, Williams SL. 2009. Accelerating loss of seagrasses across the globe threatens coastal

ecosystems. *Proceedings of the National Academy of Sciences of the United States of America* **106**:12377–12381 DOI 10.1073/pnas.0905620106.

- York PH, Smith TM, Coles RG, McKenna SA, Connolly RM, Irving AD, Jackson EL, McMahon K, Runcie JW, Sherman CDH, Sullivan BK, Trevathan-Tackett SM, Brodersen KE, Carter AB, Ewers CJ, Lavery PS, Roelfsema CM, Sinclair EA, Strydom S, Tanner JE, Van Dijk K-J, Warry FY, Waycott M, Whitehead S. 2016. Identifying knowledge gaps in seagrass research and management: an Australian perspective. *Marine Environmental Research* Epub ahead of print Jun 16 2016 DOI 10.1016/j.marenvres.2016.06.006.
- **Zeileis A, Hothorn T. 2002.** Diagnostic checking in regression relationships. *R News* **2**:7–10.