

1 **Effects of high temperature and CO₂ on intracellular DMSP in the cold-water coral *Lophelia***
2 ***pertusa***

3 Burdett, H.L.^{1,2,3*}, Carruthers, M.⁴, Donohue, P.J.C.³, Wicks, L.C.⁵, Hennige, S.J.⁵, Roberts, J.M.^{5,6,7},
4 Kamenos, N.A.^{3,4}

5

6 ¹ Scottish Oceans Institute, University of St Andrews, St Andrews, KY16 8LB, UK

7 ² Department of Earth and Environmental Sciences, University of St Andrews, St Andrews, KY16
8 9AJ, UK

9 ³ School of Geographical and Earth Sciences, University of Glasgow, Glasgow, G12 8QQ, UK

10 ⁴ School of Life Sciences, University of Glasgow, Glasgow, G12 8QQ, UK

11 ⁵ Centre for Marine Biodiversity and Biotechnology, Heriot-Watt University, Edinburgh, EH14 4AS,
12 UK

13 ⁶ Scottish Association for Marine Science, Oban, PA37 1QA, UK

14 ⁷ Center for Marine Science, University of North Carolina Wilmington, 601 S. College Road,
15 Wilmington, North Carolina, 28403-5928 USA

16

17 * Correspondence: email: hb57@st-andrews.ac.uk, tel: +44(0)1334 463 940, fax: +44(0)1334 463 949

18

19 Running title: Effect of OA on DMSP in *Lophelia*

20

21 **Abstract**

22 Significant warming and acidification of the oceans is projected to occur by the end of the
23 century. CO₂ vents, areas of upwelling and downwelling, and potential leaks from carbon capture and
24 storage facilities may also cause localised environmental changes, enhancing or depressing the effect
25 of global climate change. Cold-water coral ecosystems are threatened by future changes in carbonate
26 chemistry, yet our knowledge of the response of these corals to high temperature and high CO₂
27 conditions is limited. Dimethylsulphoniopropionate (DMSP), and its breakdown product
28 dimethylsulphide (DMS), are putative antioxidants that may be accumulated by invertebrates via their
29 food or symbionts, although recent research suggests that some invertebrates may also be able to
30 synthesise DMSP. This study provides the first information on the impact of high temperature (12°C)
31 and high CO₂ (817 ppm) on intracellular DMSP in the cold water coral *Lophelia pertusa* from the
32 Mingulay Reef Complex, Scotland (56°49'N, 07°23'W), where *in situ* environmental conditions are
33 mediated by tidally-induced downwellings. An increase in intracellular DMSP under high CO₂
34 conditions was observed, whilst water column particulate DMS+DMSP was reduced. In both high
35 temperature treatments, intracellular DMSP was similar to the control treatment, whilst dissolved
36 DMSP+DMS was not significantly different between any of the treatments. These results suggest that
37 *L. pertusa* accumulates DMSP from the surrounding water column; uptake may be up-regulated under
38 high CO₂ conditions, but mediated by high temperature. These results provide new insight into the
39 biotic control of deep-sea biogeochemistry and may impact our understanding of the global sulphur
40 cycle, and the survival of cold water corals under projected global change.

41

42 Keywords: climate change; cold-water coral; dimethylsulphide (DMS); dimethylsulphoniopropionate
43 (DMSP); global warming; ocean acidification

44

45 **Introduction**

46 Since the Industrial Revolution, atmospheric CO₂ emissions have continued to rise as a direct
47 result of industrial and agricultural growth, from pre-industrial levels of 280 parts per million
48 (Intergovernmental Panel for Climate Change, IPCC 2007), to current levels of ~390 ppm (NOAA
49 2012). Increased levels of anthropogenic CO₂ are projected to cause two major climatic changes that
50 may threaten marine ecosystems: global warming and ocean acidification (OA).

51 Over the last century, the average temperature of the surface ocean has risen by 0.74°C
52 (Hoegh-Guldberg et al. 2007; Solomon et al. 2007). Future IPCC projections suggest further
53 temperature increases by as much as 3.6°C (IPCC 2007). The world's oceans act as a sink for
54 atmospheric CO₂, with surface waters absorbing around 30% of anthropogenic CO₂ emissions (Form
55 and Riebesell 2012). Whilst this has played an important role in helping to mitigate the atmospheric
56 effects of climate change, there have been significant effects on seawater carbonate chemistry. During
57 the last 200 years, oceanic pH has dropped by 0.1 units and is projected to continue to decrease by a
58 further 0.3-0.5 units by 2100 (Caldeira and Wickett 2005) – a process known as OA. Although the
59 rate at which global surface warming and OA will occur can be modelled with a degree of accuracy in
60 the open ocean, the naturally high environmental variability of coastal regions makes projections
61 difficult and the biological impacts unclear (Duarte et al. 2013). Natural environmental variability
62 may be driven by downwelling of surface water (Findlay et al. 2013), upwelling of acidified deep
63 water (e.g. the California Current system; Feely et al. 2008; Gruber et al. 2012), the presence of CO₂
64 vents (e.g. Hall-Spencer et al. 2008) and biotic composition (e.g. Anthony et al. 2011). The
65 development of under-sea carbon capture and storage (CCS) facilities may also represent a new,
66 anthropogenic, source of environmental variability (Blackford et al. 2009). There is growing concern
67 that leaks from CCS infrastructure, thereby acting as a CO₂ point-source, may significantly affect
68 oceanic carbonate chemistry at a local to regional scale (Blackford et al. 2009). CO₂ vents, biotic
69 control, upwelling and CCS leaks may enhance or depress OA effects, necessitating a requirement to
70 understand the effects of acute and chronic changes in carbonate chemistry on marine biota.

71 Increases in oceanic CO₂ cause the carbonate saturation of the water to decrease. The
72 aragonite saturation horizon (ASH) is a naturally occurring boundary, below which the dissolution of
73 aragonite, a calcium carbonate polymorph, is favoured. Future OA projections suggest that the ASH
74 will shoal (i.e. become shallower) throughout the 21st century (Caldeira and Wickett 2005).
75 Scleractinian cold-water corals (CWCs) are aragonitic organisms that may be found in water depths of
76 1000s metres, and thus may be highly susceptible to OA. CWCs can form extensive reef frameworks
77 in fjord, continental shelf, slope and seamount settings (Roberts et al. 2009); the habitats they provide
78 are classified as Ecologically and Biologically Significant Areas by the Convention on Biological
79 Diversity (C.B.D. 2008). Today, only ~5% of CWC reefs are below the ASH (Form and Riebesell

80 2012). It is estimated that as much as 70% of the reefs could be below the ASH by 2100 (Form and
81 Riebesell 2012), threatening the survival of CWCs and the biodiversity they support (Roberts et al.
82 2009). Hence, there is a pressing need for further research on the impacts of climate change on cold-
83 water corals.

84 The impact of OA on cold-water corals is poorly understood, although it has been suggested
85 that calcification rates of cold-water corals have already declined in response to elevations in CO₂
86 since the Industrial Revolution (Maier et al. 2012). Under short-term (1-week) exposure to high-CO₂
87 (up to 981 µatm), calcification rates in the cold-water coral *Lophelia pertusa* decreased by up to 30%
88 (Form and Riebesell 2012). However, under longer (6 months) exposure to high-CO₂ (up to 984
89 µatm), the corals were able to continue to calcify beneath their tissue (Form and Riebesell 2012).
90 Thus, there is some evidence for resilience in cold-water corals against OA (McCulloch et al. 2012),
91 but studies have not yet taken into account the potentially synergistic effects of acidification and
92 increased temperature, despite recorded diel environmental variability in carbonate chemistry at the
93 Mingulay Reef CWC Complex, western Scotland (Findlay et al. 2013).

94 Research suggests that invertebrates may use algal-derived metabolites, such as
95 dimethylsulphoniopropionate (DMSP) in a similar way to algae, for example in response to
96 environmental challenges such as oxidative stress (Van Alstyne et al. 2006). A number of functions
97 have been described for DMSP, an algal secondary metabolite (Stefels 2000). DMSP is also the major
98 precursor to the climatically active gas dimethylsulphide (DMS), which has been linked to local
99 climate regulation through the production of aerosol particles and cloud formation (Charlson et al.
100 1987; Ayers and Cainey 2007). Azooxanthellate invertebrates have been observed to accumulate
101 DMSP from their diets, whilst zooxanthellate organisms acquire DMSP primarily from their
102 symbionts (Van Alstyne and Puglisi 2007), but may also be capable of synthesising DMSP internally
103 (Raina et al. 2013). It has been proposed that DMSP and its breakdown products may form an
104 'antioxidant cascade' (Sunda et al. 2002), able to scavenge a number of reactive oxygen species
105 (ROS), thus minimising oxidative damage. Recent research has shown that intracellular DMSP
106 concentrations in temperate (Burdett et al. 2012) and tropical (Burdett et al. 2013) calcifying
107 macroalgae may be elevated in response to high CO₂ / low carbonate saturation conditions.

108 The aim of this study was to assess the short-term effect of high temperature and high CO₂ on
109 the intracellular DMSP concentration of the cold-water coral *Lophelia pertusa* at the Mingulay Reef
110 Complex, western Scotland, where short-term variability in carbonate chemistry has recently been
111 recorded, and where significant changes in the carbonate chemistry are expected in the future (Findlay
112 et al. 2013). In this study, it was hypothesised that intracellular DMSP would increase in corals
113 exposed to high temperature and high CO₂, due to an increase in ROS production and a requirement
114 for antioxidant mechanisms.

115 **Material and methods**

116 **Study organism and sample collection location**

117 *Lophelia pertusa* is the predominant species of CWC in the north-east Atlantic (Cairns 1994; Fosså et al. 2002), forming biodiversity hotspots in deep-ocean ecosystems worldwide. Coral colonies were
118 collected from the Mingulay Reef Complex in the Sea of the Hebrides, west of Scotland (56°49'N,
119 07°23'W) during the 'Changing Oceans Expedition' in June 2012 on the Royal Research Ship *James*
120 *Cook* (cruise JC073) (Roberts and participants 2013). The Mingulay Reef Complex is characterised
121 by diel variability in carbonate chemistry, driven by tidally-induced downwellings of surface water
122 (Findlay et al. 2013). At the time of the experiment, the Mingulay Reef was exposed to in situ diel
123 $p\text{CO}_2$ changes of $>60 \mu\text{atm}$, equivalent to ~ 0.1 pH units (Findlay et al. 2013). Coral samples were
124 obtained using the *Holland-1* Remotely Operated Vehicle (ROV) 180 m below chart datum. Coral
125 colonies were acclimated to laboratory conditions for 2 days prior to starting the experiment in a
126 holding tank maintained at ambient seabed temperature, which was sufficient for physiological
127 stability in the control treatment, as demonstrated by Hennige et al. (2014).
128

129 **Experimental setup**

130 Coral colonies were fragmented and distributed between four abiotic treatments for 10 days,
131 following Hennige et al. (2014) (Table 1): control (9°C / 396 ± 2 ppm CO_2), high temperature (12°C /
132 396 ± 2 ppm CO_2), high CO_2 (9°C / 817 ± 3 ppm, $\sim\text{pH}$ 7.8) and high temperature+ CO_2 (12°C, 817 ± 3
133 ppm CO_2). Fragments from individual colonies were split between treatment tanks to prevent colony
134 pseudo-replication within treatments. Once time zero (T0) measurements for DMSP were taken (see
135 below), air enriched with CO_2 was initiated in the high CO_2 treatment tanks, controlled by a Wusthoff
136 Digimix system and monitored by a LiCor 820 gas analyser. Ambient salinity (~ 35.3) was maintained
137 throughout the experiment. Mesocosm water was partially (25-30%) changed every 2 days with
138 freshly collected, unfiltered, seawater, pumped from 70 m water depth, providing a food source to the
139 corals.

140 **Carbonate chemistry determination**

141 Over the course of the experiment, target pH was monitored daily using a Mettler Toledo
142 SevenGo pH meter (NBS). This data was checked independently by calculated carbonate chemistry
143 data from total alkalinity at set timepoints, and from measured CO_2 gas inputs (Licor-820) (these
144 results are presented in Table 1). Samples for total alkalinity (A_T) were collected in 40 ml EPA vials
145 and immediately poisoned with a saturated solution of mercuric chloride (8 μl). A_T was measured on a
146 Metrohm 702 SM Titrino using the open-cell potentiometric titration method on 20 ml sample
147 volumes with 0.01M HCl (repeatability: max. $\pm 0.1\%$ at Alkalinity $\sim 2200 \mu\text{mol kg}^{-1}$). All A_T samples

148 were analysed at $25 \pm 0.1^\circ\text{C}$ with temperature regulation using a water-bath (Grant OLS 200).
149 Certified Reference Materials (batch 109) from A.G. Dickson (Scripps Institution of Oceanography)
150 were used to standardize the acid at the beginning and end of each day of analysis. Carbonate
151 parameters were calculated from A_T and $p\text{CO}_2$ using CO2SYS (Pierrot et al. 2006) with dissociation
152 constants from Mehrbach et al. (1973), refit by Dickson and Millero (1987) and KSO_4 using Dickson
153 (1990).

154 **Coral intracellular DMSP**

155 Coral fragments ($n = 4\text{-}5$ per treatment from different colonies, 5-7 polyps per fragment) were
156 sampled at T0 and T10 days. Polyps (tissue and skeleton) were sealed in gas-tight borosilicate glass
157 vials (Wheaton) using Pharma-fix septa (Grace Alltech) in 0.33M sodium hydroxide (NaOH)
158 solution, to hydrolyse intracellular DMSP into DMS. Samples were stored in the dark at room
159 temperature until DMS analysis (within 2 months). Intracellular DMSP concentrations were
160 normalised to percent biomass (see below) following DMS quantification.

161 **Mesocosm water DMS/P**

162 Mesocosm water samples collected on T0 and T10 days were fixed for total dissolved
163 DMSP+DMS (DMS/P) and total particulate DMS/P. Water samples were filtered using GF/F 0.7- μm
164 depth filters (Millepore) according to Kiene and Slezak (2006). The filtrate and filter were sealed in
165 separate borosilicate glass vials (Wheaton) and crimped shut with Pharma-Fix septa (Grace Alltech) at
166 a final NaOH concentration of 0.33 M. Samples were stored in the dark at room temperature until
167 DMS analysis (within 2 months).

168 **DMS Quantification**

169 All samples were analysed using the purge-cryotrap-gas chromatography (GC) technique
170 (Turner et al. 1990) using a Shimadzu 2014 gas chromatograph at the University of Glasgow. The GC
171 was equipped with a 25 m capillary column (Restek RTx-5MS, injector and column temperature:
172 45°C , nitrogen carrier gas) and a sulphur-specific flame photometric detector (200°C). Sample
173 concentrations were calculated from DMSP standard calibration curves (DMSP standard from
174 Research Plus Inc). DMS detection limit for all samples was 0.64 ng; analytical precision was within
175 3%. Intracellular DMSP is presented as nmol g^{-1} biomass; water samples are presented as nmol L^{-1} .

176 **Percent biomass**

177 The mass of each *L. pertusa* polyp was recorded before (tissue and skeleton) and after
178 (skeleton only) storage in 0.33 M NaOH. The difference between the initial and final mass was
179 considered biomass and used to express the intracellular concentration of DMSP.

180 **Statistical Analyses**

181 To test for differences between T0 and T10 control treatments, a 2-sample t-test was used for
182 intracellular DMSP (no data transformation was required to meet test assumptions of normality and
183 homogeneity of variance) and dissolved DMS/P (x^2 data transformation). Particulate DMS/P data
184 could not be transformed to meet parametric test assumptions, so a non-parametric Mann-Whitney
185 test was used. To test for differences between treatments at T10, an ANOVA General Linear Model
186 was used for intracellular DMSP (no data transformation required) and dissolved DMS/P (x^2 data
187 transformation). Particulate DMS/P data could not be transformed to meet parametric test
188 assumptions, so a multiple-comparison Kruskal Wallis test was used. Minitab V14 was used for all
189 statistical tests.

190

191 **Results**

192 **Carbonate chemistry**

193 Table 1 presents the average ($X \pm SE$) carbonate chemistry parameters associated with each
194 mesocosm treatment over the course of the 10-day experiment. Control and high CO₂ treatments were
195 maintained at $9.06 \pm 0.02^\circ\text{C}$ and $8.90 \pm 0.11^\circ\text{C}$ respectively, whilst the high temperature and high
196 temperature+CO₂ treatments were maintained at $12.3 \pm 0.10^\circ\text{C}$ and $12.3 \pm 0.02^\circ\text{C}$ respectively (X
197 $\pm SE$, Table 1). Aragonite saturation state (Ω_{Arg}) in the control and high temperature treatments was
198 1.82 ± 0.05 and 2.01 ± 0.10 respectively ($X \pm SE$, Table 1). Ω_{Arg} in the high CO₂ treatment was 1.09
199 ± 0.01 and 1.30 ± 0.04 in the high temperature+CO₂ treatment ($X \pm SE$, Table 1). Measurements of pH
200 using the handheld meter (NBS scale) were 8.00 ± 0.02 , 7.99 ± 0.02 , 7.94 ± 0.03 and 7.90 ± 0.02 for
201 control, high temperature, high CO₂, and high temperature+CO₂ treatments respectively. The
202 relatively high pH means of the high CO₂ and high temperature+CO₂ treatments include
203 measurements following water changes with unfiltered deep seawater which was not pre-bubbled with
204 CO₂. The pH ranges for the high CO₂ and the high-temperature+CO₂ treatments were 7.76-8.07 and
205 7.75-8.03 respectively, driven by the frequent water changes (Hennige, pers. obs.). Whilst this did not
206 exactly mimic the observed diurnal variability at Mingulay (Findlay et al. 2013), the range in pH is
207 similar to that expected under projected future conditions.

208 **Intracellular DMSP in *Lophelia pertusa***

209 No significant difference in intracellular DMSP concentration was observed between T0 and
210 T10 control measurements ($t_6 = 0.12$, $P = 0.908$, Figure 1). Intracellular DMSP concentrations in the
211 T10 high CO₂ treatment ($X \pm SE = 35.36 \pm 6.96 \text{ nmol g}^{-1} \text{ biomass}$) were higher than the T10 control
212 treatment ($16.03 \pm 3.25 \text{ nmol g}^{-1} \text{ biomass}$), although this was marginally insignificant ($F_{(3,18)} = 2.56$, P

213 = 0.094, Tukey's pairwise comparison: $P = 0.077$, Figure 2). The T10 high temperature (22.52 ± 3.05
214 nmol g^{-1} biomass) and high temperature+CO₂ treatments ($20.43 \pm 7.25 \text{ nmol g}^{-1}$ biomass) were not
215 significantly different to the T10 control (Tukey's pairwise comparison: $P = 0.81$ and $P = 0.94$
216 respectively; Figure 2).

217 **Mesocosm dissolved DMS/P**

218 A significant decline in dissolved DMS/P concentrations was observed between T0 and T10 control
219 measurements ($t_4 = 6.89$, $P = 0.002$, Figure 1). In contrast, no significant difference between T10
220 treatments was observed ($F_{(3,15)} = 1.21$, $P = 0.349$, Figure 2).

221 **Mesocosm particulate DMS/P**

222 A significant decrease in particulate DMS/P concentrations was observed in the control
223 treatment from T0 to T10 ($W = 26.0$, $P = 0.030$, Figure 1). A significant difference in particulate
224 DMS/P concentrations was also observed between T10 treatments ($H_3 = 12.79$, $P = 0.005$, Figure 2):
225 particulate DMS/P in the high temperature treatment ($X \pm \text{SE} = 15.0 \pm 0.88 \text{ nmol L}^{-1}$) was
226 significantly higher than the high CO₂ ($4.68 \pm 0.17 \text{ nmol L}^{-1}$; pairwise comparison: $P = 0.002$) and
227 high temperature+CO₂ treatments ($5.19 \pm 0.54 \text{ nmol L}^{-1}$; pairwise comparison: $P = 0.005$; Figure 2).
228 Particulate DMS/P in the T10 control ($10.80 \pm 0.18 \text{ nmol L}^{-1}$) was not significantly different to any of
229 the T10 experimental treatments (Figure 2).

230

231 **Discussion**

232 **Intracellular DMSP**

233 This is the first study to assess the intracellular DMSP concentrations of *Lophelia pertusa*.
234 Samples were taken from different colonies from within the Mingulay Reef Complex, accounting for,
235 at least in part, the observed variations in DMSP concentrations. Similarly large variations in
236 intracellular DMSP concentrations between coral colonies have also been reported for tropical,
237 zooxanthellate species (Broadbent et al. 2002). Intracellular DMSP concentrations from this study
238 were similar to other invertebrate Phyla such as Porifera and Mollusca (Van Alstyne and Puglisi
239 2007), but were, surprisingly, higher than other azooxanthellate Cnidaria, which are often below
240 detection limits (Van Alstyne and Puglisi 2007; Van Alstyne et al. 2009). This highlights the potential
241 importance of cold-water corals as a store for DMSP. Given previous studies of DMSP in
242 invertebrates (e.g. Van Alstyne et al. 2006), intracellular DMSP in *L. pertusa* is likely to be
243 nutritionally derived because of *L. pertusa*'s azooxanthellate life mode. However, recent research
244 suggests that tropical corals may be able to internally synthesise DMSP, independent of zooxanthallae

245 (Raina et al. 2013); the existence of this mechanism in *L. pertusa* cannot be ruled out. Here, a modest
246 increase in intracellular DMSP in the high CO₂ treatment was observed, supporting previous research
247 of CO₂-driven up-regulation of intracellular DMSP in some phytoplankton (Arnold et al. 2012) and
248 macroalgae (Burdett et al. 2012; Burdett et al. 2013), although this has not been uniformly observed
249 and may be species-specific (Hopkins et al. 2010; Kerrison et al. 2012; Spielmeyer and Pohnert
250 2012). High CO₂ may lead to an energetic imbalance in *L. pertusa* (Hennige et al. 2014), perhaps
251 enhancing the requirement for secondary metabolites such as DMSP. However, because of the short
252 experimental period of this study, it is still difficult to predict the longer-term acclamatory responses
253 that will occur under an ocean acidification scenario.

254 High temperature had an apparently antagonistic effect on high CO₂ impacts, perhaps because
255 of an increase in feeding rate under higher temperatures (Dodds et al. 2007). However, the high
256 temperature used in this study, although above the coral's preferred temperature range in the north
257 east Atlantic (6-9 °C; Cairns 1994), has previously been observed within the Mingulay Reef Complex
258 (Davies et al. 2009). Additionally, the Mingulay Reef is known to be exposed to large seasonal and
259 diel variations in temperature (Findlay et al. 2013), so a degree of thermal tolerance was perhaps to be
260 expected.

261 **Mesocosm DMS/P**

262 DMS/P in the mesocosm water column may have been derived from algae, detritus and/or the
263 corals, but does more accurately represent in situ conditions compared to the provision of an artificial
264 food source for the corals. The method adopted here provides a total DMSP+DMS concentration;
265 different ratios in source matter and / or variations in DMSP / DMS production will have contributed
266 to the observed variations in water column DMS/P concentrations. Dissolved DMS/P did not differ
267 significantly between treatments, although a decline in dissolved DMS/P and particulate DMS/P was
268 observed from T0 to T10 in under control conditions, perhaps reflecting a change in the natural
269 planktonic composition of the water pumped from 70 m depth. Nevertheless, comparisons between
270 the control and treatment conditions may still be made at T10. In the high CO₂ and high
271 temperature+CO₂ treatments, particulate DMS/P was depleted, suggesting that accumulation by the
272 corals may have reduced mesocosm DMS/P concentrations, as has been observed in azooxanthellate
273 shallow-water cnidarians (e.g. Van Alstyne et al. 2006). The extent to which DMSP accumulation
274 may be regulated by invertebrates in response to physiological requirements (e.g. as an antioxidant) is
275 only speculative at present (Van Alstyne and Puglisi 2007), but this study suggests that *L. pertusa*
276 may up-regulate DMSP uptake under high CO₂ conditions. Further, rate-determinant studies should
277 seek to quantify DMSP uptake by *L. pertusa*.

278 The impact of high CO₂ and high temperature on the planktonic food supply for *L. pertusa* is
279 important to consider. The effect of projected environmental change on phytoplankton DMS/P

280 dynamics is likely to be highly species-specific (Wingenter et al. 2007; Vogt et al. 2008; Hopkins et
281 al. 2010; Avgoustidi et al. 2012), affecting phytoplankton bloom community composition, and
282 impacting the nutritional quality and DMSP content of the *L. pertusa* diet. This is particularly
283 important in the Mingulay Reef complex given the daily downwellings of surface water (Findlay et al.
284 2013). This, combined with an impact on coral metabolism and feeding rate, may significantly affect
285 rates of intracellular DMSP accumulation by *L. pertusa*, impacting the use of DMSP as an antioxidant
286 by the coral.

287 **Environmental implications and conclusions**

288 This study provides new information on the potential for biotic control on deep sea
289 biogeochemistry that may inform future biogeochemical studies on CWC reefs. A combination of
290 short-term (e.g. vents, up/downwelling, CCS leaks) and long-term (ocean acidification) sources of
291 high CO₂, combined with projected increases in water temperature may expose CWCs to high CO₂
292 and high temperature conditions for varying time periods. Despite the experimental limitations, this
293 study suggests that the CWC *L. pertusa* may accumulate DMSP in its tissues in response to short-
294 term, consistent exposure to high CO₂, perhaps for use as an antioxidant; further studies are, however,
295 needed to confirm this. Consequently, in areas of high coral density, DMSP accumulation may
296 remove particulate DMS/P from the surrounding water column, reducing the availability of DMS/P to
297 other organisms such as microbes, which are known to utilise DMSP and DMS as a carbon and
298 sulphur source (Green et al. 2011; Hatton et al. 2012), potentially affecting the ecosystem function of
299 local habitats. Additionally, in shallow marine environments, intracellular DMSP acts as a settlement
300 cue for a number of invertebrate larvae (Huggett et al. 2006; Steller and Cáceres-Martinez 2009). If
301 this is also the case in CWC habitats, a high CO₂-driven up-regulation of DMSP by *L. pertusa* may
302 alter the dynamics of the CWC ecosystem. Tidal downwellings connect the Mingulay Reef Complex
303 to the surface ocean. Uptake of DMSP and DMS by the corals may therefore be a, previously
304 overlooked, factor in the magnitude of surface ocean DMSP production vs sea-air DMS flux and the
305 subsequent potential for climate regulation through cloud formation.

306

307 **Acknowledgements**

308 This paper is a contribution to the UK Ocean Acidification Research Programme (Natural
309 Environment Research Council (NERC) grant NE/H017305/) and to the Marine Alliance for Science
310 & Technology Scotland (MASTS). This research was conducted whilst HLB was initially in receipt of
311 NERC studentship funding (NE/H525303/1) and ultimately a MASTS Research Fellowship, PJCD
312 was in receipt of a MASTS PhD studentship, NAK was in receipt of Royal Society of Edinburgh /
313 Scottish Government Fellowship (RES 48704/1) and SJH was in receipt of a NERC Independent

314 Research Fellowship (NE/K009028/1). SJH, LCW and JMR acknowledge support from Heriot-Watt
315 University's Environment & Climate Change Theme. Many thanks extend to the captain and crew of
316 the RRS *James Cook* (cruise 073) for assistance at sea. The data from this study are freely available
317 from the British Oceanographic Data Centre (Burdett et al. 2014).

318

319 References

- 320 Anthony KRN, A. Kleypas J, Gattuso J-P (2011) Coral reefs modify their seawater carbon chemistry
321 – implications for impacts of ocean acidification. *Global Change Biology* 17: 3655-3666 doi
322 10.1111/j.1365-2486.2011.02510.x
- 323 Arnold HE, Kerrison P, Steinke M (2012) Interacting effects of ocean acidification and warming on
324 growth and DMS-production in the haptophyte coccolithophore *Emiliana huxleyi*. *Global*
325 *Change Biology*: n/a-n/a doi 10.1111/gcb.12105
- 326 Avgoustidi V, Nightingale PD, Joint I, Steinke M, Turner S, Hopkins FE, Liss PS (2012) Decreased
327 marine dimethyl sulfide production under elevated CO₂ levels in mesocosm and in vitro
328 studies. *Environmental Chemistry* 9: 399-404
- 329 Ayers GP, Cainey JM (2007) The CLAW hypothesis: a review of the major developments.
330 *Environmental Chemistry* 4: 366-374 doi doi:10.1071/EN07080
- 331 Blackford J, Jones N, Proctor R, Holt J, Widdicombe S, Lowe D, Rees A (2009) An initial assessment
332 of the potential environmental impact of CO₂ escape from marine carbon capture and
333 storage systems. *Proceedings of the Institution of Mechanical Engineers, Part A: Journal of*
334 *Power and Energy* 223: 269-280
- 335 Broadbent AD, Jones GB, Jones RJ (2002) DMSP in corals and benthic algae from the Great Barrier
336 Reef. *Estuar Coast Shelf Sci* 55: 547-555 doi DOI 10.1006/ecss.2002.1021
- 337 Burdett HL, Aloisio E, Calosi P, Findlay HS, Widdicombe S, Hatton AD, Kamenos NA (2012) The
338 effect of chronic and acute low pH on the intracellular DMSP production and epithelial cell
339 morphology of red coralline algae. *Mar Biol Res* 8: 756-763
- 340 Burdett HL, Carruthers M, Donohue PJC, Wicks LC, Hennige SJ, Roberts JM, Kamenos NA (2014)
341 Impacts of increased CO₂ and temperature upon DMSP production in the cold-water coral,
342 *Lophelia pertusa*, from short-term experiments carried out on cruise D366/7. *British*
343 *Oceanographic Data Centre - Natural Environment Research Council, UK* doi: 10/rf4
- 344 Burdett HL, Donohue PJC, Hatton AD, Alwany MA, Kamenos NA (2013) Spatiotemporal Variability
345 of Dimethylsulphoniopropionate on a Fringing Coral Reef: The Role of Reefal Carbonate
346 Chemistry and Environmental Variability. *PLoS ONE* 8: e64651
- 347 C.B.D. (2008) COP 9 Decision IX/20: Marine and coastal biodiversity
- 348 Cairns SD (1994) *Scleractina of the temperate North Pacific*. Smithsonian Contributions to Zoology.
349 Smithsonian Institution Press, Washington D.C.
- 350 Caldeira K, Wickett ME (2005) Ocean model predictions of chemistry changes from carbon dioxide
351 emissions to the atmosphere and ocean. *J Geophys Res* 110: C09S04
- 352 Charlson RJ, Lovelock JE, Andreae MO, Warren SG (1987) Oceanic phytoplankton, atmospheric
353 sulphur, cloud albedo and climate. *Nature* 326: 655-661
- 354 Dickson AG (1990) Standard potential of the reaction: $\text{AgCl}(s) + 1/2\text{H}_2(g) = \text{Ag}(s) + \text{HCl}(aq)$, and and
355 the standard acidity constant of the ion HSO_4^- in synthetic sea water from 273.15 to 318.15
356 K. *The Journal of Chemical Thermodynamics* 22: 113-127
- 357 Dickson AG, Millero FJ (1987) A comparison of the equilibrium constants for the dissociation of
358 carbonic acid in seawater media. *Deep Sea Research Part A Oceanographic Research*
359 *Papers* 34: 1733-1743

360 Dodds LA, Roberts JM, Taylor AC, Marubini F (2007) Metabolic tolerance of the cold-water coral
361 *Lophelia pertusa* (Scleractinia) to temperature and dissolved oxygen change. *J Exp Mar*
362 *Biol Ecol* 349: 205-214

363 Duarte C, Hendriks I, Moore T, Olsen Y, Steckbauer A, Ramajo L, Carstensen J, Trotter J, McCulloch
364 M (2013) Is Ocean Acidification an Open-Ocean Syndrome? Understanding Anthropogenic
365 Impacts on Seawater pH. *Estuaries and Coasts* 36: 221-236 doi 10.1007/s12237-013-9594-3

366 Feely RA, Sabine CL, Hernandez-Ayon JM, Ianson D, Hales B (2008) Evidence for Upwelling of
367 Corrosive "Acidified" Water onto the Continental Shelf. *Science* 320: 1490-1492 doi
368 10.1126/science.1155676

369 Findlay HS, Artioli Y, Moreno Navas J, Hennige SJ, Wicks LC, Huvenne VAI, Woodward EMS,
370 Roberts JM (2013) Tidal downwelling and implications for the carbon biogeochemistry of
371 cold-water corals in relation to future ocean acidification and warming. *Global Change*
372 *Biology*: n/a-n/a doi 10.1111/gcb.12256

373 Form AU, Riebesell U (2012) Acclimation to ocean acidification during long-term CO₂ exposure in
374 the cold-water coral *Lophelia pertusa*. *Global Change Biology* 18: 843-853 doi
375 10.1111/j.1365-2486.2011.02583.x

376 Fosså JH, Mortensen PB, Furevik DM (2002) The deep-water coral *Lophelia pertusa* in Norwegian
377 waters: distribution and fishery impacts. *Hydrobiologia* 471: 1-12

378 Green DH, Shenoy DM, Hart MC, Hatton AD (2011) DMS oxidation coupled to biomass production
379 by a marine Flavobacterium. *Appl Environ Microbiol* 77: 3137-3140

380 Gruber N, Hauri C, Lachkar Z, Loher D, Frölicher TL, Plattner G-K (2012) Rapid Progression of Ocean
381 Acidification in the California Current System. *Science* 337: 220-223 doi
382 10.1126/science.1216773

383 Hall-Spencer JM, Rodolfo-Metalpa R, Martin S, Ransome E, Fine M, Turner SM, Rowley SJ, Tedesco
384 D, Buia M-C (2008) Volcanic carbon dioxide vents show ecosystem effects of ocean
385 acidification. *Nature* 454: 96-99

386 Hatton A, Shenoy D, Hart M, Mogg A, Green D (2012) Metabolism of DMSP, DMS and DMSO by the
387 cultivable bacterial community associated with the DMSP-producing dinoflagellate
388 *Scrippsiella trochoidea*. *Biogeochemistry* 110: 131-146 doi 10.1007/s10533-012-9702-7

389 Hennige SJ, Wicks LC, Kamenos NA, Bakker DCE, Findlay HS, Dumousseaud C, Roberts JM (2014)
390 Short-term metabolic and growth responses of the cold-water coral *Lophelia pertusa* to
391 ocean acidification. *Deep Sea Research Part II: Topical Studies in Oceanography* 99: 27-35
392 doi <http://dx.doi.org/10.1016/j.dsr2.2013.07.005>

393 Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD, Sale PF,
394 Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Inglesias-Prieto R, Muthiga N, Bradbury RH,
395 Dubi A, Hatziolos ME (2007) Coral Reefs Under Rapid Climate Change and Ocean
396 Acidification. *Science* 318: 1737-1742

397 Hopkins FE, Turner SM, Nightingale PD, Steinke M, Bakker D, Liss PS (2010) Ocean acidification and
398 marine trace gas emissions. *Proceedings of the Academy of Natural Sciences of*
399 *Philadelphia* 107: 760-765

400 Huggett M, Williamson J, de Nys R, Kjelleberg S, Steinberg P (2006) Larval settlement of the
401 common Australian sea urchin *Heliocidaris erythrogramma* in response to bacteria from
402 the surface of coralline algae. *Oecologia* 149: 604-619 doi 10.1007/s00442-006-0470-8

403 IPCC (2007) Summary for Policymakers. In: Solomon S, Qin D, Manning M, Chen Z, Marquis M,
404 Averyt KB, M.Tignor, Miller HL (eds) *Climate Change 2007: The Physical Science Basis*
405 *Contribution of Working Group I to the Fourth Assessment Report of the*
406 *Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, UK,
407 New York, USA

408 Kerrison P, Suggett D, Hepburn L, Steinke M (2012) Effect of elevated pCO₂ on the production of
409 dimethylsulphoniopropionate (DMSP) and dimethylsulphide (DMS) in two species of *Ulva*
410 (Chlorophyceae). *Biogeochemistry* 110: 5-16 doi 10.1007/s10533-012-9707-2

411 Kiene RP, Slezak D (2006) Low dissolved DMSP concentrations in seawater revealed by small-
412 volume gravity filtration and dialysis sampling. *Limnology and Oceanography: Methods* 4:
413 80-95

414 Maier C, Watremez P, Taviani M, Weinbauer MG, Gattuso JP (2012) Calcification rates and the
415 effect of ocean acidification on Mediterranean cold-water corals. *Proceedings of the Royal*
416 *Society B: Biological Sciences* 279: 1716-1723 doi 10.1098/rspb.2011.1763

417 McCulloch M, Trotter J, Montagna P, Falter J, Dunbar R, Freiwald A, Försterra G, López Correa M,
418 Maier C, Rüggeberg A, Taviani M (2012) Resilience of cold-water scleractinian corals to
419 ocean acidification: Boron isotopic systematics of pH and saturation state up-regulation.
420 *Geochim Cosmochim Acta* 87: 21-34 doi <http://dx.doi.org/10.1016/j.gca.2012.03.027>

421 Mehrbach C, Culbertson CH, Hawley JE, Pytkowicz RM (1973) Measurement of the Apparent
422 Dissociation Constants of Carbonic Acid in Seawater at Atmospheric Pressure. *Limnol*
423 *Oceanogr* 18: 897-907

424 NOAA (2012) Trends in atmospheric carbon dioxide

425 Pierrot D, Lewis E, Wallace DWR (2006) CO2SYS DOS Program developed for CO₂ system
426 calculations. ORNL/CDIAC-105. Carbon Dioxide Information analysis Center. Oak Ridge
427 National Laboratory, US Department of Energy, Oak Ridge, TN

428 Raina J-B, Tapiolas DM, Foret S, Lutz A, Abrego D, Ceh J, Seneca FO, Clode PL, Bourne DG, Willis BL,
429 Motti CA (2013) DMSP biosynthesis by an animal and its role in coral thermal stress
430 response. *Nature* 502: 677-680 doi 10.1038/nature12677

431 Roberts J, participants (2013) Changing Oceans Expedition 2012. RRS *James Cook* 073 Cruise
432 Report.

433 Roberts JM, Wheeler A, Freiwald A, Cairns SD (2009) Cold-water Corals: The Biology and Geology
434 of Deep-sea Coral Habitats. Cambridge University Press, Cambridge

435 Solomon S, Qin D, Manning M, Chen Z, Marquis M, Averyt KB, Tignor M, Miller HL (2007)
436 Contribution of Working Group I to the Fourth Assessment Report of the
437 Intergovernmental Panel on Climate Change, 2007 Cambridge University Press, Cambridge,
438 UK and New York, USA

439 Spielmeyer A, Pohnert G (2012) Influence of temperature and elevated carbon dioxide on the
440 production of dimethylsulfoniopropionate and glycine betaine by marine phytoplankton.
441 *Marine Environmental Research* 73: 62-69

442 Stefels J (2000) Physiological aspects of the production and conversion of DMSP in marine algae
443 and higher plants. *J Sea Res* 43: 183-197

444 Steller D, Cáceres-Martinez C (2009) Coralline algal rhodoliths enhance larval settlement and early
445 growth of the Pacific calico scallop *Argopecten ventricosus*. *Mar Ecol Prog Ser* 396: 49-60
446 doi 10.3354/meps08261

447 Sunda W, Kieber DJ, Kiene RP, Huntsman S (2002) An antioxidant function for DMSP and DMS in
448 marine algae. *Nature* 418: 317-320

449 Turner SM, Malin G, Bagander LE, Leck C (1990) Interlaboratory Calibration and Sample Analysis of
450 Dimethyl Sulfide in Water. *Marine Chemistry* 29: 47-62

451 Van Alstyne K, Dominique V, Muller-Parker G (2009) Is dimethylsulfoniopropionate (DMSP)
452 produced by the symbionts or the host in an anemone-zooxanthella symbiosis? *Coral*
453 *Reefs* 28: 167-176

454 Van Alstyne K, Puglisi M (2007) DMSP in marine macroalgae and macroinvertebrates: Distribution,
455 function, and ecological impacts. *Aquatic Sciences* 69: 394-402

456 Van Alstyne K, Schupp P, Slattery M (2006) The distribution of dimethylsulfoniopropionate in
457 tropical Pacific coral reef invertebrates. *Coral Reefs* 25: 321-327 doi 10.1007/s00338-006-
458 0114-9

459 Vogt M, Steinke M, Turner S, Paulino A, Meyerhöfer M, Riebesell U, LeQuéré C, Liss P (2008)
460 Dynamics of dimethylsulphoniopropionate and dimethylsulphide under different CO₂

461 concentrations during a mesocosm experiment. *Biogeosciences* 5: 407-419 doi 10.5194/bg-
462 5-407-2008

463 Wingenter OW, Haase KB, Zeigler M, Blake DR, Sherwood RF, Sive BC, Paulino A, Thyrhaug R,
464 Larsen A, Schulz K, Meyerhofer M, Reibesell U (2007) Unexpected consequences of
465 increasing CO₂ and ocean acidity on marine production of DMS and CH₂ClI: Potential
466 climate impacts. *Geophysical Research Letters* 34: L05710

467

468

469

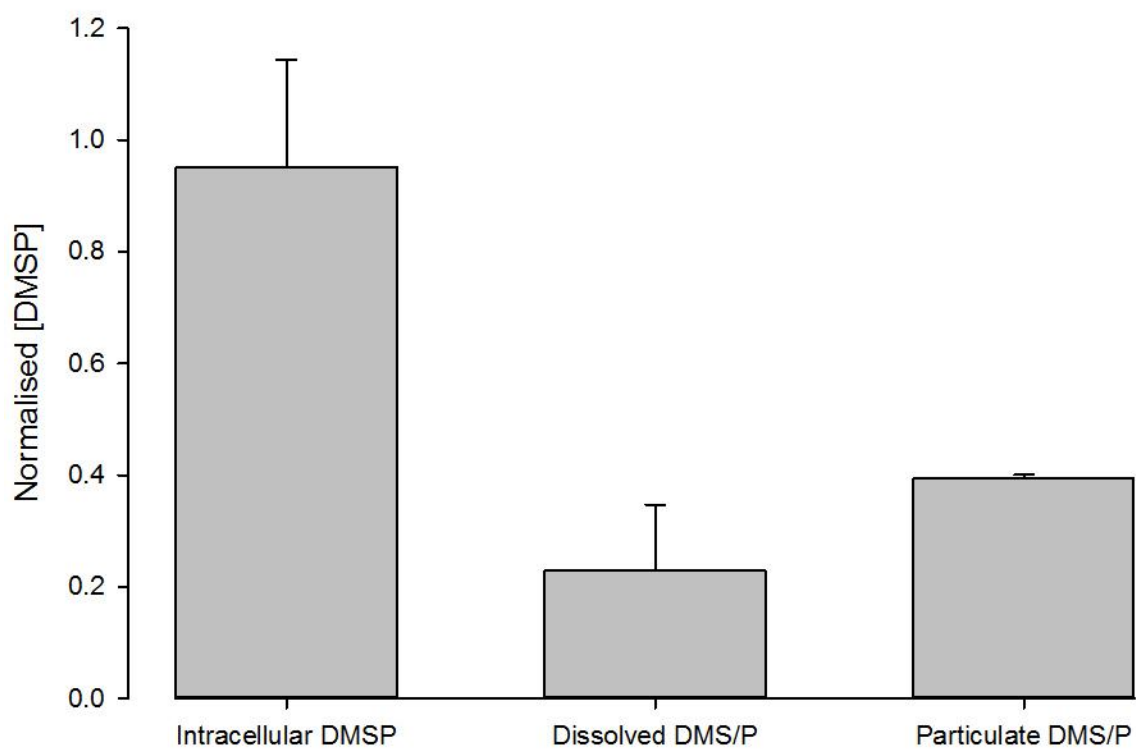
470 **Table 1.** Mean (\pm SE) values of measured temperature, salinity, injected CO₂ and total alkalinity, and
 471 derived dissolved inorganic carbon species (TCO₂), pH (Total scale) and aragonite saturation state
 472 (Ω Arg) for control, high temperature, high CO₂ and high temperature+CO₂ treatments.

| Parameter | Treatment | | | |
|---|-----------------|------------------|----------------------|----------------------------------|
| | Control | High temperature | High CO ₂ | High temperature+CO ₂ |
| Temperature (°C) | 9.06 \pm 0.02 | 12.30 \pm 0.10 | 8.90 \pm 0.11 | 12.30 \pm 0.02 |
| Salinity | 35.3 \pm 0.02 | 35.4 \pm 0.03 | 35.2 \pm 0.08 | 35.4 \pm 0.05 |
| Alkalinity (μ mol kg ⁻¹) | 2197 \pm 34.9 | 2170 \pm 57.3 | 2299 \pm 8.52 | 2277 \pm 27.0 |
| TCO ₂ (μ mol kg ⁻¹) | 2031 \pm 30.5 | 1986 \pm 49.5 | 2227 \pm 7.86 | 2189 \pm 24.8 |
| pH _T | 8.03 \pm 0.01 | 8.02 \pm 0.01 | 7.76 \pm 0.00 | 7.76 \pm 0.00 |
| pCO ₂ (ppm) | 396 \pm 2.08 | 396 \pm 2.08 | 817 \pm 2.49 | 817 \pm 2.49 |
| Ω Arg | 1.82 \pm 0.05 | 2.01 \pm 0.10 | 1.09 \pm 0.01 | 1.30 \pm 0.04 |

473

474

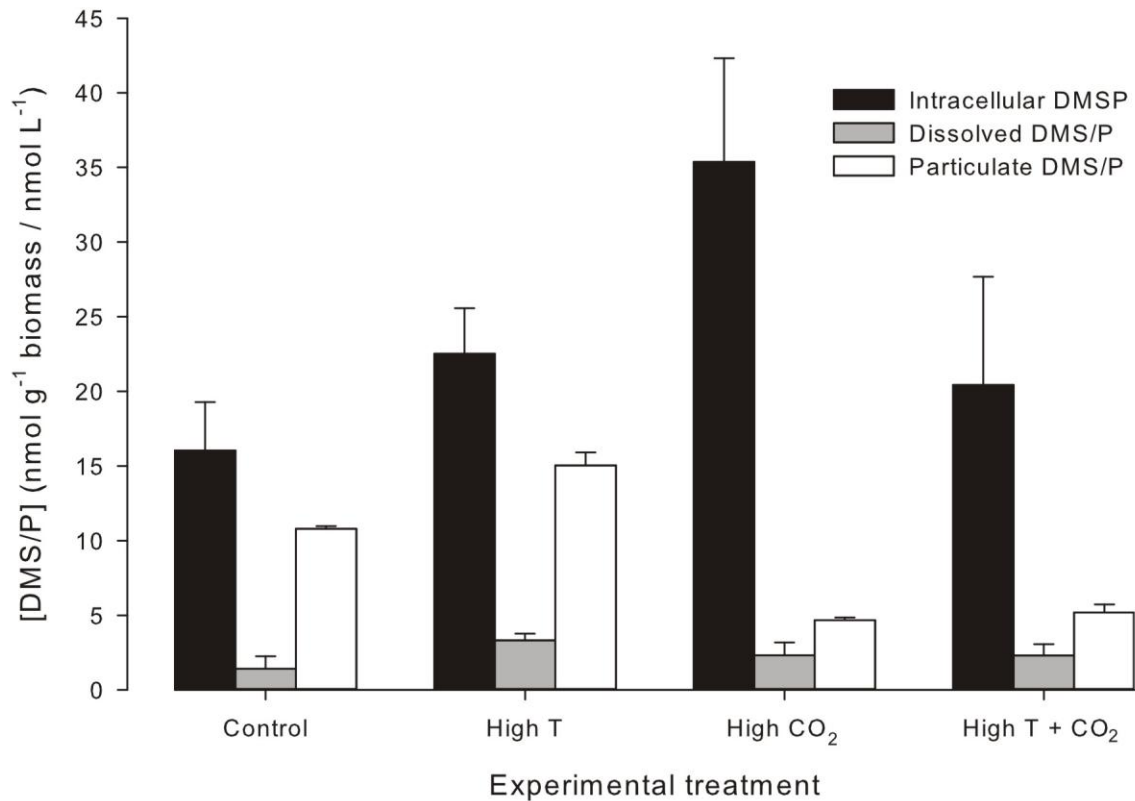
475



476

477 **Fig.1** *Lophelia pertusa* intracellular DMSP (n = 5) and dissolved and particulate DMS/P
478 concentrations (n = 4) in the control treatment at the end of the 10-day experiment, normalised against
479 control measurements at the beginning of the experiment. See Table 1 for a summary of the
480 experimental conditions. Data presented as mean ± SE

481



482
 483 **Fig.2** DMSP concentrations following 10-days incubation at control, high temperature, high CO₂ and
 484 high temperature+CO₂ conditions. *Lophelia pertusa* intracellular DMSP (normalised to g of coral
 485 tissue biomass; n = 5 except in high temperature+CO₂ treatment where n = 4) and dissolved (n = 4)
 486 and particulate (n = 4) DMS/P concentrations are presented. See Table 1 for a summary of the
 487 experimental conditions. Data presented as mean ± SE