



UNIVERSITÀ DEGLI STUDI DI TORINO

This is the author's final version of the contribution published as:

Marucco, Arianna; Pellegrino, Francesco; Oliaro-Bosso, Simonetta; Maurino, Valter; Martra, Gianmario; Fenoglio, Ivana

Indoor illumination: A possible pitfall in toxicological assessment of photo-active nanomaterials

in Journal of Photochemistry and Photobiology A: Chemistry, 2018, January, 350, 23-31, Available online 18 September 2017.

The publisher's version is available at: 10.1016/j.jphotochem.2017.08.072

When citing, please refer to the published version.

Link to this full text: http://hdl.handle.net/2318/317106

This full text was downloaded from iris-Aperto: https://iris.unito.it/

1	Indoor illumination: a possible pitfall in toxicological
2	assessment of photo-active nanomaterials
3	
4	Arianna Marucco ^{a,b} , Francesco Pellegrino ^a , Simonetta Oliaro-Bosso ^c , Valter Maurino ^a ,
5	Gianmario Martra ^{a,b} , Ivana Fenoglio ^{a,b*}
6	
7	^a Department of Chemistry, University of Torino, via P. Giuria 7, 10125 Torino, Italy.
8	^b NIS – Nanostructured Interfaces and Surfaces and 'G. Scansetti'' Interdepartmental Centre
9	for Studies on Asbestos and other Toxic Particulate, University of Torino, Italy.
10	^c Department of Drug Science and Technology, via P. Giuria 7, 10125 Torino, University of
11	Torino. Italy.
12	
13	* corresponding author: Ivana Fenoglio, Department of Chemistry, University of Torino, via
14	P. Giuria 7, 10125 Torino, Italy. e.mail ivana.fenoglio@unito.it. Phone +39 11 6707506.
15	
16	Abstract
17	Standardization of the experimental protocols used in the hazard assessment of nanomaterials
18	(NMs) is strongly required to reduce inconsistency among data deriving by different
19	laboratories. The parameters that are known to modify the toxic response of cells to NMs are
20	in fact higher than for soluble toxicants. Among them illumination, that may induce activation
21	of some semiconducting NMs, has been poorly investigated.
22	The present study, conducted within the FP7 EU project SETNanoMetro, has been designed to
23	assess the effect of indoor illumination on the oxidative potential and dispersion degree of nano-
24	TiO ₂ . The generation of Reactive Oxygen Species (ROS) by four nanometric anatase or rutile-

25	anatase TiO ₂ specimens under ordinary laboratory illumination has been evaluated by means of
26	Electronic Paramagnetic Resonance (EPR) spectroscopy, while their ability to damage DNA
27	has been measured by agarose gel electrophoresis using plasmid DNA as model. The effect of
28	illumination on nanoparticles dispersion has been evaluated by Dynamic Light Scattering
29	(DLS). The results show the occurrence of photo-activation of TiO_2 under indoor illumination
30	that leads to the generation of ROS and slight plasmid DNA damage. Furthermore, significant
31	differences in the amount of ROS generated were found for small variation of the intensity of
32	the illumination. A small effect on the size distribution of TiO ₂ agglomerates in water was
33	observed.
34	The present findings suggest that illumination should be included among the parameters that
35	have to be controlled during toxicological assessment of photo-active nanomaterials.
36	
37	Keywords
38	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination.
38 39	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination.
38 39 40	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination.
38 39 40 41	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the
 38 39 40 41 42 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on
 38 39 40 41 42 43 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method
 38 39 40 41 42 43 44 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials
 38 39 40 41 42 43 44 45 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials (NMs) standardization is a particularly relevant issue. Numerous specific and non-specific
 38 39 40 41 42 43 44 45 46 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials (NMs) standardization is a particularly relevant issue. Numerous specific and non-specific factors have been shown to influence the results of toxicological testing [3], such as: i) the
 38 39 40 41 42 43 44 45 46 47 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials (NMs) standardization is a particularly relevant issue. Numerous specific and non-specific factors have been shown to influence the results of toxicological testing [3], such as: i) the degree of dispersion of the NM, that varies depending upon the media used [4,5,6], ii) the real
 38 39 40 41 42 43 44 45 46 47 48 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials (NMs) standardization is a particularly relevant issue. Numerous specific and non-specific factors have been shown to influence the results of toxicological testing [3], such as: i) the degree of dispersion of the NM, that varies depending upon the media used [4,5,6], ii) the real dose, that, oppositely to molecular substances may not correspond to the nominal one [7,8], iii)
 38 39 40 41 42 43 44 45 46 47 48 49 	Nanomaterials; reactive oxygen species; particle dispersion; toxicological testing; illumination. 1. Introduction The knowledge of the hazard is a fundamental pre-requisite to reduce the risk associated to the exposure to chemicals. The current European regulation (REACH), places responsibility on industry to provide safety information on the substances. As consequence, test method standardization for hazard assessment is strongly needed [1, 2]. In the case of nanomaterials (NMs) standardization is a particularly relevant issue. Numerous specific and non-specific factors have been shown to influence the results of toxicological testing [3], such as: i) the degree of dispersion of the NM, that varies depending upon the media used [4,5,6], ii) the real dose, that, oppositely to molecular substances may not correspond to the nominal one [7,8], iii) the presence of contaminants in the materials like bacterial lipopolysaccharides (LPS) [9] or

absorbance/fluorescence of the material [3,11,12]. Much less explored is the effect of 51 52 illumination during the preparation of samples and NM exposure in toxicological testing. Except in the case in which a specific illumination is necessary due to the kind of endpoint 53 evaluated, incubation of cells in *in vitro* tests is performed in the dark. On the other hand, the 54 preparation of the NMs and the administration to cells is performed in laboratories illuminated 55 by artificial or natural light. During these steps, photo-activation of semiconductors materials 56 like ZnO, CeO₂, NiO or TiO₂ may occur since indoor natural illumination and some artificial 57 light (e.g. halogen lamps) contains UV radiation. 58

Among them titanium dioxide (TiO₂) and TiO₂-based materials are the most widespread NMs 59 [13,14], being used for several purposes, e.g. as UV blockers in sunscreens and plastics [15, 60 16]. TiO₂ is a powerful photo-catalyst. When illuminated with UV light it generates at its 61 surface a high amount of reactive species, a property that finds application in several fields, like 62 63 in water and air remediation [17] or in the production of self-cleaning coatings and textiles [18]. The adsorption of photons with energies higher or equal to the TiO_2 band gap (>3.2 eV for 64 anatase) results in electrons to be excited in the conduction band (e⁻CB) leading to the formation 65 of a positive hole in the valence band (h^+_{VB}) . These charge carriers can recombine each other 66 or migrate at the surface where they react with electron donors or acceptors that diffuse close 67 68 to the surface [19]. For example, by reacting with water and oxygen, hydroxyl radicals (HO⁻), superoxide radicals (O_2^{-}) , singlet oxygen $({}^1O_2)$, hydroperoxyl radicals ('OOH) and hydrogen 69 peroxide (H₂O₂) are formed. The generation of such oxygenated radicals and molecules, 70 commonly called Reactive Oxygen Species (ROS), affects the ability of TiO₂ to interact with 71 cells by increasing its oxidative potential, i.e. the ability to induce an oxidative burst [20, 21]. 72 The role of particle-derived ROS in the in the photo-toxicity of TiO₂ is well established [22, 73 74 23]. On the other hand, the toxicity of non-illuminated TiO_2 is not expected to be related to them [24]. Nevertheless, several studies performed in the absence of specific illumination 75 reported TiO₂-induced effects related to the occurrence of oxidative burst [25, 26]. Whether it 76

is a consequence of ROS generated by light-activated TiO₂ or of cell-derived ROS is not clear
 since in most of the studies the illumination condition used during NMs handling is not
 described.

Another well-known property of TiO₂ is the superhydrophilicity: under irradiation with UV 80 light the abundance of hydrophilic groups at the surface of TiO₂ (Ti-OH) increases, an effect 81 that is reversed in the dark [27]. Superhydrophilicity may affect the agglomeration degree of 82 TiO₂ suspensions in water. In fact, particle agglomeration may occur in colloidal suspensions 83 when particles exhibit a low osmotic repulsion. In this case attractive van der Waals forces and 84 entropy driven surface dehydration prevail leading to agglomeration [28]. This largely depends 85 86 on the thickness of the Stern layer around particles [29] that in turn depends upon particles surface chemistry, in particular the abundance and type of charged groups. 87

Indeed, a light-induced disaggregation of TiO₂ nanoparticles under UV light was previously
reported [30].

The present study is aimed to assess the effect of indoor illumination on the oxidative potential and dispersion degree of nano-TiO₂. Four samples of fully characterized nano-TiO₂ in the anatase or anatase-rutile forms, the most photo-active ones, has been selected and analyzed for their ability to generate ROS in different illumination conditions by using a set of EPR-based tests previously proposed as integrated protocol for the assessment of biological-relevant photoactivity of TiO₂[31]. The ability to damage DNA was also tested. Finally, the effect on the NM dispersion, evaluated by means of Dynamic Light Scattering (DLS) analysis was investigated.

97

98 **2. Methods**

2.1 TiO₂ samples. Four types of titania NMs were considered, three commercial materials (i.e.
P25 by Evonik Industries, Germany; SX001 by Solaronix, Swizerland; and PC105 by Cristal,
Saudi Arabia) and one lab-made TiO₂ prepared via hydrothermal synthesis, and then coded as
UT001. Details of preparation, structural and morphological characterization of the specimens

are in ref. [32]. In brief, UT001 was obtained by forced hydrolysis of an aqueous solution of 103 $Ti(TeoaH)_2$ complex (Teoa = triethanolamine; initial pH 10), carried out by hydrothermal 104 treatment at 453 K for 90 h. Before the use, each material was suspended in water and then 105 processed according to the following procedure in order to remove organic and inorganic 106 impurities adsorbed onto TiO_2 NPs: i) dialysis against ultrapure water (MilliQ, Millipore) using 107 a Spectra/Por dialysis membrane tubing (MWCO 8-12 kD or MWCO 12-14 kD); final pH of 108 the permeated liquid in the 5-6 range, Cl⁻ and $SO_4^{2-} < 1$ ppm (by ion chromatography); ii) freeze-109 drying; iii) re-suspension in milli-Q water; iv) irradiation for 48 hours of the suspension in 110 contact with air, added of 10 ml of H₂O₂ (30%), under UV light using a medium pressure 111 mercury lamp (emission max at 360 nm), ca. 50 W/m² in the range 290-400 nm; followed by 112 dialysis and free-drying as steps i) and ii). Step iv) ensures a complete photo-degradation of 113 organic impurities adsorbed onto TiO₂ NPs that can change their surface properties and 114 115 reactivity of TiO₂. H₂O₂ is used as electron scavenger to speed up impurities degradation.

2.2 Surface Area Measurements. The specific surface area (SSA_{BET}) of the powders was
 measured by adsorption of N₂ at 77 K, applying the BET model for the analysis of results.

2.3 X-ray diffraction (XRD). X-ray diffraction (XRD) pattern of the powders were recorded
with an Analytical X'Pert Pro equipped with an X'Celerator detector powder diffractometer
using Cu Ka radiation generated at 40 kV and 40 mA. The instrument was configured with 1/2°
divergence and receiving slits. A quartz sample holder was used. The 2θ range was from 20° to
80° with a step size (°2θ) of 0.05 and a counting time of 3 s.

123 2.4 Morphological Characterization. TEM images were obtained with a Jeol 3010 124 instrument, operated at 300 kV. For the observation, powders were contacted in dry form with 125 standard Cu grids coated with a lacey carbon film, and then introduced in the microscope. To 126 evaluate the presence of aggregates the samples were also analyzed by Dynamic Light 127 Scattering in 200 mM ammonia solution, in order to maximize the particles electrostatic 128 repulsion, after 30 min sonication and adjusting the concentration depending on the sample 129 characteristics. The Dynamic Light Scattering system used was an ALV (Langen Germany), 130 NIBS model (non invasive backscattering) with fixed scattering angle (173°). Through the 131 Stokes-Einstein equation the hydrodynamic radius $r_{\rm H}$ of the agglomerates/aggregates were 132 obtained.

2.5 Diffuse reflectance UV-Vis spectroscopy. The optical behaviour of the powders in the 133 UV-Vis range was investigated by electronic absorption spectroscopy in the diffuse reflectance 134 mode. Spectra were acquired with a Cary 5000 instrument (Varian), equipped with an 135 integrating sphere coated with Spectalon[®], also used as reference. In order to avoid side effects 136 137 due to differences in particle packing, the powder cell provided by Varian was used, allowing pressing a sample toward the optically pure quartz window constituting the front part of the 138 cell. Proper amounts of powders were used, resulting in layers of ca. 3 mm in thickness, thus 139 reaching the usual condition for correct measurements in the diffuse reflectance mode [33]. 140

141 **2.6 Illumination conditions and light irradiance measurements.**

142 The irradiance in the visible and UV regions was measured with a photo-radiometer (Delta Ohm S. r. L., Padova, Italy) under natural indoor light (windows closed) and artificial 143 illumination (halogen lamp). For the sake of comparison, the irradiance of outdoor natural light 144 145 was also measured (Table 1). Measurements were made twice a day for one week in the month of September (Latitude: 45°04'13" Ν 2016 146 Longitude: 7°41'12" E). No UVC radiation is expected to be present in solar light and therefore 147 was not measured. 148

Experiments were performed under reduced illumination (shielded light) set up under a hood by shielding the glasses, or under standardized light obtained by using a 500 W Hg/Xe lamp (Oriel Instruments) equipped with an IR water filter to avoid the overheating of the suspensions with or without a 400 nm cut-off filter. The presence of the 400 nm cut-off filter leads a radiation that contains a fraction of UVA/B of intensity intermediate between outdoor and indoor (Table 1). Finally experiments were also performed in a dark room with a red led as unique source of light. In this case, each sample was weighted and then kept in the dark room for 24h beforeexperiments, in order to exclude any activation of the powders.

157

Light source				
	Visible 1050-400 nm	UVA 400-315 nm	UVB 315-280 nm	Total UVA/UVB (calculated)
Hg/Xe lamp	795 ± 73	432 ± 60	836 ± 78	1268
Outdoor	315 ± 119	3.30 ± 1.19	0.28 ± 0.13	3.6
Halogen lamp	7.30 ± 1.50	63.9x10 ⁻³ ± 5.0x10 ⁻³	7.9x10 ⁻³ ± 1.0x10 ⁻³	71.8x10 ⁻³
Natural indoor light	3.30 ± 1.40	25.6x10 ⁻³ ± 8.0x10 ⁻³	0.97 x10 ⁻³ ± 0.3x10 ⁻³	26.6x10 ⁻³
Hg/Xe lamp + filter	560 ± 58	0.164 ± 0.040	3.5 x10 ⁻³ ± 3.0 x10 ⁻³	0.168
Shielded light	0.20 ± 0.06	0.70x10 ⁻³ ± 1.0x10 ⁻³	0.75 x10 ⁻³ ± 0.07	1.4x10 ⁻³
Dark	18.7x10 ⁻³ ± 0.30x10 ⁻³	0.10x10 ⁻³ ± 0.08x10 ⁻³	0.60x10 ⁻³ ± 0.2x10 ⁻³	0.7x10 ⁻³

158 Table 1. Light irradiance at the different illumination conditions

159

160

2.7 Generation of free radicals and singlet oxygen. All experiments were performed in
ultrapure MilliQ water (Millipore, Billerica, MA). Amounts of powders corresponding the same
exposed surface area (1.4 m²) calculated on the basis of the SSA_{BET} was used for all
experiments (P25 25mg; PC105 16mg; UT001 30mg: SX001 15mg). The powders were
transferred in a 2.5 ml quartz vial and generation of the different ROS was monitored by adding
the following solutions:

167 1) Total reactivity: 2 ml of a 50µM solution of TEMPONE-H (1-hydroxy-2,2,6,6-tetramethyl-

168 4-oxo-piperidine, Enzo Life Sciences Inc., Farmingdale, New York, US) in water;

2) Oxidative reactivity: 0.5 ml of a solution of DMPO (5,5-Dimethyl-1-pyrroline-N-oxide,
Enzo Life Sciences Inc., Farmingdale, New York, US) 88 mM, and sodium formate 1M in
phosphate buffer saline (pH 7.5, 0.005M);

3) Singlet oxygen generation: 2 ml of a 50 mM solution of 4-oxo-TMP (2,2,6,6-tetramethyl-4piperidone, Sigma-Aldrich, Saint Louis, Missouri, US) in phosphate buffer saline (pH 7.4,
0.01M)

The suspensions were exposed to the different illumination conditions for 60 or 40 minutes and the generation of radical species monitored by Electron Spin Resonance (EPR) spectroscopy (Miniscope 100 EPR spectrometer, Magnettech, Berlin, Germany) on aliquots of the suspensions withdrawn with a glass capillary each 10 minutes.

Instrument settings: microwave power 7 mW, modulation amplitude 1G, scan time 80s, two
scans. The negative controls were, in all experiments, the solutions illuminated in the same
conditions as the samples. All experiments were repeated at least three times.

The amount of radical generated was evaluated by building a calibration curve with the stable
free radicals 4-oxo-TEMPO (or TEMPONE, 4-oxo-2,2,6,6-tetramethylpiperidine-1-oxyl, Enzo
Life Sciences Inc., Farmingdale, New York, US) in water in the concentrations range 50 - 0.12
μM.

186 **2.8 Generation of hydrogen peroxide.** An amount of powder corresponding to an exposed surface area of 1.4 m² was suspended in 2 ml of water in a quartz vial and exposed to the chosen 187 illumination conditions. The powder was removed by filtration (cellulose acetate, 0.20 µm). 188 The concentration of hydrogen peroxide on the supernatant was evaluated by using the method 189 reported by Mottola et al. [34]. 50 mg of leucocrystal violet (LCV) was dissolved in 80 ml of 190 0.5%(v/v) HCl and diluted to 100 ml with the same solution. A buffer solution was made by 191 mixing equal volume of 2M sodium acetate and 2M of acetic acid and adjusting the pH to 4.5 192 with acetic acid. 193

1 ml of LCV solution was added to a 1 ml of the supernatant. 4 ml of buffer and 0.5 ml of 194 195 peroxidase (type I from horseradish) (1 mg/mL) was added and the solution diluted to 10 ml with water. The absorbance was measured after 10 minutes at of the sample at 596 nm against 196 197 a reference prepared in the same manner but with no powder (Kontron Instruments Inc., Everett, MA). The concentration of hydrogen peroxide was determined by building a calibration curve. 198 199 **2.9 Direct plasmid DNA damage.** Plasmids are convenient model systems to study direct DNA 200 damage because their sizes are well defined, the quantification of their single breaks (SSBs) by gel electrophoresis is relatively easy and accurate, the chemical environment of the DNA can 201 be precisely controlled, and there is no biological repair processes [35]. Here pYES2 plasmid 202 203 DNA (Invitrogen, Italy) was used as a model. The damage was quantified in terms of single (SSB) and double (DSB) breaks in the DNA strand. Strand breaks were detected by agarose (1 204 %) gel electrophoresis which separates the three forms of DNA molecules, supercoiled DNA 205 206 (undamaged plasmid); open circular DNA (resulting from single-strand breaks); linear DNA (a 207 product of double-strand breaks). Experiments were performed in a quartz vial with 0.2 mg of powder suspended in 30 µl of MilliQ water and then vortexed. To this suspension 5 µl of DNA 208 209 solution (about 50 ng/µl) were added and then exposed for 20 minutes to the different illumination conditions. As control, DNA was exposed to the corresponding illumination 210 211 condition, for the same time in the absence of any powder in order to exclude a direct damage to this molecule. After the exposure time (20 min) the suspension was centrifuged (15000 g) 212 and the supernatant was used for gel electrophoresis. DNA bands were stained and visualized 213 214 with ethidium bromide (Promega, Italy).

Controls of the different forms of plasmid DNA were obtained by digesting the supercoiled
DNA with EcoRI enzyme in the presence or in the absence of ethidium bromide [23].

2.10 Hydrodynamic diameter. The hydrodynamic diameter was evaluated by dynamic light
 scattering (DLS) (Zetasizer Nano-ZS, Malvern Instruments, Worcestershire, U.K., detection
 limits 1 nm-6 μm) in ultrapure water or in a 0.05 wt% solution of bovine serum albumin (BSA,

Sigma-Aldrich, Saint Louis, Missouri, US) by using a dispersion protocol adapted from the EU-220 FP7 project NanoGenoTox deliverable 3 (http://www.nanogenotox.eu). Briefly, a 2.56 mg/ml 221 222 stock dispersion was prepared by pre-wetting powder in 0.5 vol% ethanol (96% purity) followed by dispersion in 0.05 wt% BSA and sonicated for 35 minutes with a probe sonicator 223 224 (100 W, 40% amplitude, 20 kHz, 3 mm titania probe, Sonoplus, Bandelin, Berlin, Germany). The time of sonication and the amplitude of the sonicator power were set-up to deliver a pre-225 determined acoustic power according to the method developed within the EU-FP7 project 226 227 NANoReg. DLS analysis was started after 10 minutes of incubation in the various illumination conditions. The results are the mean of three independent measurements each consisting in 228 consecutive 10 runs on the same vial. 229

3. Results

3.1 Physico-chemical characterization of the TiO₂ samples

- 233 The XRD patterns of the four selected materials show that all materials are pure anatase, except
- for P25 which is a mixture of anatase/rutile (Table 2 and SI).
- 235

	Crystalline phaseª	Impurities	Specific surface area (m²/g) ^b	Primary particle size (nm) ^c	Hydrodynamic diameter (nm, NH₃ 0.2 M) ^d
P25	Anatase 80% Rutile 20%	-	55	30	36
PC105	Anatase >99%	sulphates	86	23	630 (aggregates)
UT001	Anatase >99%	carbonates/ carboxylates	47	33	34
SX001	Anatase	carbonates/ carboxylates	93	19	82 (aggregates)

236Table 2. Main physico-chemical features of the TiO2 samples

- 237 ^a XRD; ^b BET; ^c TEM; ^dDLS
- 238

UT001 exhibits a quite regular bi-pyramidal shaped nanoparticles with regular borders [32]
whereas P25, PC105 [36] and SX001 [37] are characterized by less regular profiles (Figure S1
in SI), in agreement with the presence of a significant fraction of surface terminations different
from {101}previously facets (the most stable ones), as probed by IR spectroscopy of adsorbed
CO [37].

The trend exhibited by the specific surface area (SSA) is in qualitative agreement with the size of primary nanoparticles (the smaller the size, the larger the SSA). Nevertheless, only UT001 and P25 nanoparticles attained a mono-dispersion when suspended in a proper aqueous medium, whereas even in the best dispersion condition attained the other two TiO₂ powders 248

249

exhibited hydrodynamic diameters larger than primary particles, indicating they are constituted by agglomerates of nanoparticles, quite huge in size for PC105 [38].

All samples appeared, as expected, opaque to UV radiation and transparent to visible light 250 (Figure S2 in SI). The Kubelka-Munk vs. wavelength spectra exhibited the typical absorption 251 edge due to the valence-to-conduction band transition [39]. As expected, the absorption edge 252 of P25 (anatase:rutile \approx 80:20 by weight) is located at longer wavelength with respect the pure 253 254 anatase materials, because of the narrower inter-band energy gap of rutile [40]. In all cases the onset of the absorption is located in the high energy part of the visible range, due to the coupling 255 256 of photon absorption with phonon emission, one of the two results of the indirect character of the inter-band transition [41,42]. Nevertheless, an additional localized absorption seems to be 257 present in the visible range of the spectrum of PC105, which could be due to localized charged-258 transfer absorptions related to the presence on the surface of these TiO₂ nanoparticles of 259 sulphate groups. 260

261

262 **3.2 Effect of UV radiation intensity of TiO₂ photo-reactivity**

Preliminary experiments conducted in normal indoor light unexpectedly showed a significant photo-activation of all powders. However, the amount of radicals generated was highly variable during the day and the seasons, because of the different intensity of the light (Table 1). The activation of the TiO_2 powders was also monitored by gradually shielding the laboratory natural light down to the minimal amount of visible light allowing to operate, and by measuring the reactivity of the powder toward sodium formate at fixed irradiance values .



270

Figure 1. Effect of light intensity on TiO₂ activation. EPR signals recorded after 60 minutes of
incubation of A) P25; B) UT001; C) PC105; D) SX001 in a buffered solution (PBS, 0.005M. pH 7.5)
containing 1M sodium formate and 88mM DMPO. The irradiance values (UVA) measured during the
experiments are reported on the graphs.

275

The EPR signal intensity (Figure 1), that is proportional to the powder reactivity, was found largely dependent on the light irradiance for all the samples, despite the differences in the UVA radiation intensity was very low. At a value of UVA of 0.7 mW/m², correspondent to a total UVA+UVB 1.4 mW/m² (Table 1) the lowest signal intensity was obtained for all the samples. This last condition, referred here as "shielded light" can be considered the condition most widely applicable in practice. For this reason, was chosen for the subsequent experiments.

3.3 TiO₂ photo-activation in shielded light

- A set of tests were used to measure the overall reactivity of TiO_2 that may induce cell damage,
- including all ROS species generated by TiO₂ and the oxidative/reductive processes that follow

the direct reaction of molecules with the surface charge carriers [43]. This was achieved by 285 EPR spectroscopy using three different probes, sodium formate, the hydroxylamine 286 TEMPONE-H and the piperidone 4-oxo-TMP. Hydrogen peroxide generation was evaluated 287 by a spectrophotometric method as described in the Method Section. The experiments were 288 performed under the shielded light condition described above and the data compared with those 289 obtained by using the filtered Hg/Xe lamp (positive control) or in the dark (negative control). 290 In Figure 2 the amount of radicals generated in the three illumination conditions, at the last time 291 292 point considered in the kinetic, are reported. The full kinetics of generation are reported in the SI. A substantial reactivity toward the three probes was observed following irradiation with the 293 filtered Hg/Xe lamp, while no hydrogen peroxide generation was detected in any conditions 294 (not reported). Some differences among samples was observed: UT001 and SX001 were the 295 most reactive toward sodium formate (Figures 2 and S5 in SI), UT001 and P25 appeared the 296 297 most reactive toward TEMPONE-H (Figures 2 and S4 in SI), while the most active in generating singlet oxygen was SX001 (Figures 2 and S6 in SI). An increase of the amount of 298 299 ROS generated with time was observed for all samples in all tests (SI).



300

306

Figure 2. ROS generation by the TiO₂ samples in different illumination conditions. A) radicals
 generated in the presence of sodium formate (oxidative reactivity); B) radicals generated in the presence
 of TEMPONE-H (total reactivity); C) generation of singlet oxygen. The data are expressed as amount
 of radicals generated per unit surface area of the powder, at the last point considered in the kinetic (see
 SI). Illumination conditions are indicated in each panel.

The amount of ROS generated in the shielded light condition was negligible when measured with sodium formate or 4-oxo-TMP. However, the samples, and in particular UT001 and P25, was still active toward TEMPONE-H, albeit the amount of ROS generated was one order of magnitude lower than those generated with the filtered Hg/Xe lamp. Differently to what previously reported [44], in the dark a negligible reactivity was observed. This was because in the present case the powders were kept for 24h in the dark, thus suggesting a possible role of pre-illumination.

- 315
- 316

3.4 TiO₂-induced damage to plasmid DNA in shielded light

The reactivity was evaluated by incubating the powders with SC-pDNA in the various illumination conditions and by measuring the DNA damage by agarose electrophoresis (Figure 3). The induction of DNA strand breaks was evaluated by the conversion of the supercoiled form (SC) to open circular (OC) and linear (L) forms.

321



Figure 3. Effect of illumination on TiO₂ -induced damage to double stranded supercoiled plasmid DNA. The plasmid DNA was exposed for 20 min to the TiO₂ samples in different illumination conditions (irr: filtered Hg/Xe lamp, sl: shielded light, d: dark). Damage was evaluated as capability to induce the formation of, open circular DNA (OC) and linear DNA(L) from native supercoiled double stranded DNA (SC). (M) marker; (DNA irr) DNA irradiated by the filtered Hg/Xe lamp without powders; (DNA EcoRI) DNA digested with EcoRI enzyme; (DNA EcoRI+EtBr) DNA digested with EcoRI enzyme + ethidium bromide.

330

322

As expected, when illuminated with the filtered Hg/Xe lamp, all the TiO₂ samples caused a clear damage to DNA: in fact, an increase in the intensity ratio between the bands correspondent to the open circular (OC) plasmid DNA and supercoiled double stranded DNA (SC) was observed for all samples with respect to plasmid DNA irradiated in absence of powder and the DNA treated with the powders in the dark. This effect is a clearly consequence of a ROS dependent or independent oxidative damage directly induced by activated TiO₂. No bands correspondent to the linear form (L) of DNA was observed. When exposed to shielded light, a very slight increase of the OC/SC intensity band ratio was observed with all samples exceptthan for PC105.

340 **3.5 Light-induced aggregation/disaggregation**

To evaluate the possible effect of illumination on the agglomerate size distribution of TiO₂ 341 UT001 was chosen. This sample was the best candidate for DLS analysis since it is composed 342 by particles having a narrow size distribution. The mean hydrodynamic diameter of the powder 343 and the polydispersion index (PDI) were measured in both water and in a 0.05% bovine serum 344 albumin (BSA) solution after sonication following a standardized protocol as described in the 345 method section. Measurements were performed after exposing the suspension for 10 minutes 346 in the dark, indoor light and filtered Hg/Xe lamp. A further condition, i.e. illuminating with the 347 full range of visible, UVA and UVB radiations (Hg/Xe lamp without filter) was also used to 348 induce the highest activation possible of the powders. 349



352 353

354

355 356 **Figure 4. Effect of illumination on TiO**₂ **dispersions.** Panels A and B: hydrodynamic diameters of UT001 in A) water and B) 0.05% BSA solution, under different illumination conditions: dark (a), indoor light (b), filtered Hg/Xe lamp (c), Hg/Xe lamp (d). Panel C: Z average values (bars) and the polydispersion indexes (PdI) (points) of UT001 in water (black) and in BSA solution (gray).

In water, the powders appeared organized in agglomerated (Figure 4). No significant variations 357 in mean hydrodynamic diameter were found depending by illumination, except for a moderated 358 shift when the suspension was irradiated with the Hg/Xe lamp. However, a significant decrease 359 of the PDI was observed in all illumination conditions suggesting that disaggregation occurred 360 at some extent. When dispersed in the 0.05% BSA the TiO₂ suspension appeared more 361 uniformly dispersed (lower polydispersion index) in the dark than in water. This was expected 362 since proteins can act as surfactants by adsorbing at the surface and increasing the repulsion 363 among particles. Still, the particles appeared agglomerated, with a small fraction of 364 monodisperse particles. When illuminated, a shift of the mean hydrodynamic diameter toward 365 higher diameters was unexpectedly observed. This effect was particularly relevant after 366 illumination with the UV/vis light. In this conditions, the Z-average value was three time higher 367 than those observed in the dark. 368

369

4. Discussion

The data presented herein demonstrate that, in spite of the very low amount of UV light in normal indoor natural illumination ($15.94 \pm 4.8 \text{ J/m}^2$ during 10 minutes), photo-activation of TiO₂ occurs, leading possible photo-induced effects on cells during toxicological testing. Since photo-activation is largely dependent upon the intensity of UVA/B radiation, illumination may be considered a possible a source of variability of the toxicological data obtained in different laboratories.

4.1 Identification of the ROS generated in shielded indoor illumination.

TiO₂ is able to generate several different ROS. Photo-generated electrons may reduce oxygen to superoxide radicals (equation 1) while holes oxidize water leading to the generation of the highly reactive hydroxyl radicals (equation 2).

381 $O_2 + e^- \rightarrow O_2^{--} (1)$

382 $HO^{-} + h^{+}_{vb} \rightarrow HO^{-}(2)$

Hydroxyl and superoxide radicals may further react to generate secondary species that are hydroperoxyl radicals, the conjugated acid of the superoxide anion (equation 3), and hydrogen peroxide (equation 4). Hydrogen peroxide may further react with conduction band electrons generating hydroxyl radicals (equation 5).

387
$$O_2^{-} + H_2O \leftrightarrows HO_2^{-} + H^+(3)$$

388
$$O_2^{-} + e^- + 2H^+ \rightarrow H_2O_2(4)$$

389 $H_2O_2 + e^- \rightarrow HO^- + HO^- (5)$

Singlet oxygen $({}^{1}O_{2})$ is also generated by a mechanism still under discussion [45, 46].

Formate ions are sensitive probes for the evaluation of the oxidative reactivity of TiO_2 [47], since they are able to react with both photo-generated holes and hydroxyl radicals to form carboxylate radicals, but not with superoxide radicals. Oppositely, TEMPONE-H measures the total reactivity of TiO_2 , being able to reacts with all species, included superoxide radicals, forming the stable radical specie TEMPONE.

In the shielded light condition used, the reactivity of the powders toward sodium formate was very low (Figure 5A). At the same time no singlet oxygen (Figure 5C) or H_2O_2 generation (data not shown) was observed. On the other hand, small amount of TEMPONE radicals were detected for SX001 and UT001in the presence of TEMPONE-H (Figure 5B), likely generated following reaction of the probe with superoxide radicals. The ability of TiO₂ to stabilize superoxide radicals by coordination with Ti⁴⁺ ions exposed at the surface [48] may account for their presence in the reaction system. Alternatively, the higher reactivity of TiO₂ toward 403 TEMPONE-H by respect to formate may be due to the scavenging of the photo-generated 404 electrons by the probe that inhibits the recombination of the charge carriers [49].

The occurrence of photo-activation suggests the possible capability of the powders to induce 405 406 oxidative damage also in the presence of very low amount of light. Among the various possible targets of oxidative damage DNA is the most relevant under a toxicological point of view, since 407 it may be related to cell death, mutation or cancer. Nucleic acids are particularly sensible to 408 ROS [50] that may induce nucleosides oxidation, inter-strand cross-links and strand breaks 409 formation [51]. The ROS most involved are hydroxyl radicals and singlet oxygen, while 410 superoxide radicals are known to be inert toward biomolecules [51]. However, they generate 411 412 other reactive species through reactions 3-5 or in biological environment through Haber-Weiss cycle [52] and therefore it may indirectly induce DNA damage. 413

The tendency of the molecules to get close enough to the surface to react with both short-living 414 415 ROS species or directly with charge carriers depends upon several factors like diffusion rate and ability to bind to it. For this reason, it is not possible to directly transfer the reactivity of 416 417 TiO₂ toward the probes used in EPR experiments to biomolecules. Therefore, the powders were further tested for their ability to induce strand breaks to DNA by a direct mechanism (not 418 mediated by cells stimulation) using supercoiled plasmid double-stranded DNA (SC-pDNA) as 419 model (Figure 3). In shielded light, a very small increase in the intensity ratio between the band 420 correspondent to the open circular plasmid DNA and supercoiled double stranded DNA was 421 observed for all samples except than for PC105. However, this effect was low if compared with 422 those observed with the positive control (filtered Hg/Xe lamp), and likely irrelevant in cells, 423 where natural antioxidant systems are present. On the other hand, we cannot exclude the 424 generation of punctual defects like oxidized nucleosides or cross-links. Further experiments 425 will be necessary to evaluate the relevance of the present findings in cells. 426

427 **4.2 Variability among anatase samples**

The analyzed samples are all uncoated and characterized by the anatase form, the most reactive one. However, they exhibit significant differences in physico-chemical features (Table 2) that may reproduce the variability encountered among the TiO₂ NM commercially available.

431 Albeit all samples are characterized by particles having a predominance of exposed {101} facets and exhibiting higher energy terminations, in the case of UT001 facets are quite regular, 432 whilst other high energy surface terminations, i.e. exposing Ti and O sites with a high 433 coordinative unsaturation level, are present on SX001 [37]. Both UT001 and SX001 have 434 carbonate and carboxylate contaminants, mainly in the bulk, while SX001 has sulphate groups 435 as surface contaminants. The samples differ also for the morphology: UT001 and P25 are 436 composed by single particles, PC105 is actually in the form of quite large aggregates while 437 SX001 is constituted by small aggregates of primary particles. Finally, P25 contains also rutile 438 nanoparticles. 439

Focusing on the optical properties of these materials, differences in surface texture and the occurrence of a limited agglomeration seemed do not result in significant difference in the absorption spectra of UT001 and SX001. Conversely, sulphation and large aggregation appeared related to an enhanced absorption of PC105 in the high energy visible range. Finally, as expected, the presence of rutile in P25 resulted in a shift of the inter-band transition edge toward longer wavelength [39].

The ability to generate ROS by the four samples examined was quantitatively and qualitatively similar. However, some differences were observed. In particular, UT001 and SX001 appeared overall more reactive on a surface area unit basis by respect to the other two samples, in the test carried out using TEMPONE-H under shielded illumination conditions (Figure 2). Such differences are not due to adsorption in the visible range, since the only samples having a small adsorption out of the UV range is PC105, but likely to the presence in the bulk and at the surface of carbonate/carboxylate species [36].

The effect on plasmid DNA was similar for all samples except for PC105 that exhibited a lower reactivity, in agreement with the lower capability to generate ROS. These data suggest that variability in term of photo-activity may be found depending upon small variation in morphological and surface properties that may be further amplified by the different surface area of the TiO₂ samples.

458 **4.3 Effect of illumination on TiO₂ dispersions.**

459 As discussed in the introduction, illumination was previously reported to affect the agglomeration degree of TiO₂ in water [30]. This is a very important issue since the size of 460 agglomerates is known to modulate both the real dose experienced by cells, due to differences 461 462 in sedimentation rates, and the cell uptake [7]. In the present case, little changes in mean size were observed in all illumination conditions when the powders were dispersed in water (figure 463 4); however the size distribution of aggregates appeared narrower when illuminated than in the 464 dark, as suggested by the lower polydispersion index, indicating that the photo-activation may 465 affect the dispersion degree of TiO₂ powders. On the other hand, when a protein was added to 466 467 the system, a clear agglomeration was observed at the more extreme illumination conditions. This effect is relevant since it may occur in cell media, where proteins are generally added as 468 nutrients for cells. To elucidate the reasons of this behaviour is out of the scope of this report 469 470 and will be the object of further investigation. However, we may speculate that the photoactivation induces conformational changes to the absorbed protein that in turn leads to particle 471 agglomeration. 472

473

474 **5.** Conclusions

In conclusion, the data presented herein indicate that the intensity of illumination during sample
preparation and exposure is an important parameter in *in vitro* testing of photo-active
semiconducting NMs. It therefore has to be controlled and accurately reported. In fact, the lack
of this information may limit the inter-laboratory comparability of toxicological data.

Albeit operating in a dark room appears to be the best condition during NMs handling, based on our results for pure anatase or anatase-rutile samples a shielded light, correspondent to a maximum of total UV irradiance of 1.4 mW/m², may be suggested to minimize the effects related to the photo-activation of the samples. Cellular studies are in progress to validate this conclusion. Note however, that the present finding apply only to NMs having optical properties similar to TiO₂, and not to all photo-active NM, like for example the forms of TiO₂ purposely designed to absorb in the visible region by doping or adsorption of dyes.

486

487 Acknowledgments

This project has received funding from the European Union, Seventh Programme (FP7/2007-488 2013) under the project "Shape-engineered TiO₂ nanoparticles for metrology of functional 489 properties: setting design rules from material synthesis to nanostructured devices 490 491 "(SETNanoMetro), grant agreement No. 604577 and under the project "A common European approach to the regulatory testing of nanomaterials" (NANoREG), grant agreement No 492 493 310584. The Authors are grateful to Alessandra Buffa and Dr. Magda Mocanu that performed 494 the set-up of the illumination conditions and of DLS measurements as part of their postgraduated thesis. 495

496

497 **References**

[1] B. Fadeel, V. Kagan, H. Krug, A. Shvedova, M. Svartengren, L. Tran, L. Wiklund, There's
plenty of room at the forum: Potential risks and safety assessment of engineered nanomaterials,

- 500 Nanotoxicology. 1 (2007) 73-84.
- 501 [2] A. Pietroiusti, Health implications of engineered nanomaterials, Nanoscale. 4 (2012) 1231502 1247.
- 503 [3] V. Stone, H. Johnston, RPF. Schins, Development of in vitro systems for nanotoxicology:
- methodological considerations, Crit Rev Toxicol. 39 (2009) 613–626.

505	[4] Opinion	on: Risk	Assessme	nt of Products	of Nanotec	hnologies,	Scientific Comm	nittee or
506	Emerging	and	Newly	Identified	Health	Risks	(SCENIHR),	2009.
507	http://ec.euro	opa.eu/he	ealth/scient	ific_committee	es/emerging/	/opinions/s	scenihr_opinions_	en.htm
508	#nano							

- 509 [5] H. Bouwmeester, I. Lynch, H.J. Marvin, K.A. Dawson, M. Berges, D. Braguer, H.J. Byrne,
- A. Casey, G. Chambers, M.J. Clift, G. Elia, T.F. Fernandes, L.B. Fjellsbø, P. Hatto, L.
 Juillerat, C. Klein, WG. Kreyling, L.C. Nicke, M. Riediker, V. Stone, Minimal analytical
 characterization of engineered nanomaterials needed for hazard assessment in biological
 matrices, Nanotoxicology. 5 (2011) 1-11.
- [6] F. Catalano, G. Alberto, P. Ivanchenko, G. Dovbenko, and G. Martra, Effect of silica surface
 properties on the formation of multilayer or submonolayer protein hard corona: albumin
 adsorption on pyrolytic and colloidal SiO₂ nanoparticles, J. Phys. Chem. C. 119 (2015) 2649326505.
- [7] J. Ponti, R. Colognato, H. Rauscher, S. Gioria, F. Broggi, F. Franchini, C. Pascual, G.
 Giudetti, F. Rossi, Colony forming efficiency and microscopy analysis of multi-wall carbon
 nanotubes cell interaction, Toxicol Lett. 197 (2010) 29-37.
- 521 [8] D. Lison, G. Vietti, S. van der Brule, Paracelsus in nanotoxicology, Particle and Fibre
 522 Toxicology. 11 (2014) 35.
- 523 [9] M.G. Bianchi, M. Allegri, A.L. Costa, M. Blosi, D. Gardini, C. Del Pivo, A. Prina-Mello,
- L. Di Cristo, O. Bussolati, E. Bergamaschi, Titanium dioxide nanoparticles enhance
 macrophage activation by LPS through a TLR4-dependent intracellular pathway, Toxicol Res.
 4 (2015) 385-398.
- 527 [10] E. Aldieri, I. Fenoglio, F. Cesano, E. Gazzano, G. Gulino, D. Scarano, A. Attanasio, G.
- 528 Mazzucco, D. Ghigo, B. Fubini, The role of iron impurities in the toxic effects exerted by short
- 529 MWCNT in murine alveolar macrophages, J Toxicol Environ Health Part A. 76 (2012) 1056-
- 530 1071.

531	[11] F. Schrurs, D. Lison, Focusing the research efforts, Nature Nanotech. 7 (2012) 546-548.
532	[12] A. Marucco, F. Catalano, I. Fenoglio, F. Turci, G. Martra, F. Fubini, Possible chemical
533	source of discrepancy between in vitro and in vivo tests in nanotoxicology caused by strong
534	adsorption of buffer components, Chemical research in toxicology. 28 (2015) 87-91.
535	[13] F. Piccinno, F. Gottschalk, S. Seeger, B. Nowack, Industrial production quantities and uses
536	of ten engineered nanomaterials in Europe and in the world, Nanopart Res. 14 (2012) 1109-
537	1118.
538	[14] OECD, Guidance document 116 on the conduct and design of chronic toxicity
539	and carcinogenicity studies, supporting test guidelines 451,452 and 453. 2nd edition
540	Series on Testing and Assessment No. 116 ENV/JM/MONO (2011)
541	[15] Official Journal of the European Union, Regulation (EC) No.1223/2009 of the
542	european parliament and of the council of 30 November 2009 on cosmetic products.
543	[16] N. Serpone, D. Dondi, A. Albini, Inorganic and organic UV filters: Their role and efficacy
544	in sunscreens and suncare products, Inorg Chim Acta. 360 (2007) 794-802.
545	[17] A.G. Agrios, P. Pichat, State of the art and perspectives on materials and applications of
546	photocatalysis over TiO ₂ , J appl electrochem. 35 (2005) 655-663.
547	[18] S. Ortelli, M. Blosi, S. Albonetti, A. Vaccari, M. Dodi, A.L. Costa, TiO ₂ based nano
548	photocatalysis immobilized on cellulose substrates, Photoch Photobio A. 276 (2014) 58-64.
549	[19] M. Chiesa, M.C. Paganini, S. Livraghi, E. Giamello, Charge trapping in TiO ₂ polymorphs
550	as seen by Electron Paramagnetic Resonance spectroscopy, Phys. Chem. Chem. Phys. 15
551	(2013) 9435-9447.
552	[20] T. Finkel, Redox-dependent signal transduction, Febs let. 476 (2000) 52-54.
553	[21] P. Møller, N.R. Jacobsen, J.K. Folkmann, P.H. Danielsen, L. Mikkelsen, J.G.
554	Hemmingsen, L.K. Vesterdal, L. Forchhammer, H. Wallin, S. Loft, Role of oxidative damage
555	in toxicity of particulates, Free Radical Res. 44 (2010) 1-46.

- [22] H. Ma, A. Brennan, S.A. Diamond, Photocatalytic reactive oxygen species production and
 phototoxicity of titanium dioxide nanoparticles are dependent on the solar ultraviolet radiation
 spectrum, Environ Toxicol Chem. 31 (2012) 2099-2107.
- [23] I. Fenoglio, J. Ponti, E. Alloa, M. Ghiazza, I. Corazzari, R. Capomaccio, D. Rembges, S.
- Oliaro-Bosso, F. Rossi, Singlet oxygen plays a key role in the toxicity and DNA damage of
 nanometric TiO₂ to human keratinocytes, Nanoscale. 5 (2013) 6567-6576.
- 562 [24] S. Dalai, S. Pakrashi, R.S.S. Kumar, N. Chandrasekaran, A. Mukherjee, A comparative
- 563 cytotoxicity study of TiO_2 nanoparticles under light and dark conditions at low exposure 564 concentrations, Toxicol Res. 1 (2012) 116-130.
- 565 [25] International Agency for research on cancer (IARC): Carbon black, Titanium dioxide and
- talc. in: IARC Monographs on the evaluation of carcinogenic risk to humans. vol 93, Lyon,2010.
- 568 [26] S.T. Larsen, P. Jackson, S.S. Poulsen, M. Levin, K.A. Jensen, H. Wallin, G.D. Nielsen, I.
- K. Koponen, Airway irritation, inflammation, and toxicity in mice following inhalation of metal
 oxide nanoparticles, Nanotoxicology. 10 (2016) 1254-1262.
- [27] S.H. Wang, T.K. Chen, K.K. Rao, M.S. Wong, Nanocolumnar TiO₂ thin films uniquely
 incorporated with carbon for visible light photocatalysis, Appl Catal B-Environ. 76 (2007) 328334.
- 574 [28] J. Israelachvili, Intermolecular and surface surfaces, second ed., Academic Press, London,
 575 1991.
- 576 [29] D.J. Shaw, Introduction to Collid and Surface Chemistry, fourth ed., Butterworth-577 Heinemann, 1992.
- 578 [30] S.W. Bennet, D. Zhou, R. Mielke, A. Keller, Photoinduced disaggregation of
 579 TiO₂ nanoparticles enables transdermal penetration, Plos one. 7 (2012) 1-7.

580	[31] A. Marucco, E. Gazzano, D. Ghigo, E. Enrico, I. Fenoglio, Fibrinogen enhances the
581	inflammatory response of alveolar macrophages to TiO2, SiO2 and carbon nanomaterials,
582	Nanotoxicology. 10 (2016) 1-9.

- 583 [32] C. Deiana, M. Minella, G. Tabacchi, V. Maurino, E. Fois, G. Martra, Surface features of
- TiO₂ nanoparticles: combination modes of adsorbed CO probe the stepping of (101) facets,
- 585 Phys Chem Chem Phys. 15 (2013) 307-315.
- [33] G. Kortum, Reflectance Spectroscopy: Principles, Methods, Applications, Springer Verlag, New York, 1969
- 588 [34] H.A. Mottola, B.E. Simpson, G. Gorin, Absorbimetric determination of hydrogen
- peroxide submicrogram amounts with leouco crystal violet and peroxidase as catalyst, Anal
 Chem. 42 (1970) 410-411
- [35] M.R. Gual, F.M. Milan, A. Deppman, P.R.P. Coelho, Study of DNA damage with a new
- 592 system for irradiation of samples in a nuclear reactor, Appl Radiat Isotopes. 69 (2011) 373-376.
- 593 [36] C.L. Bianchi, S. Gatto, C. Pirola, A. Naldoni, A. Di Michele, G. Cerrato, V. Crocellà, V.
- 594 Capucci, Photocatalytic degradation of acetone, acetaldehyde and toluene in gas-phase:
- 595 Comparison between nano and micro-sized TiO_2 , Appl Catal B-Environ. 146 (2014) 123-130.
- [37] C. Deiana, E. Fois, G. Martra, S. Narbey, F. Pellegrino, G. Tabacchi, On the simple
 complexity of carbon monoxide on oxide surfaces: facet-specific donation and back donation
 effects revealed on TiO₂ anatase nanoparticles, Chem Phys Chem. 17 (2016) 1956–1960.
- [38] F. Pellegrino, L. Pellutiè, F. Sordello, C. Minero, E. Ortel, V.-D. Hodoroaba, V. Maurino,
 Influence of agglomeration and aggregation on the photocatalytic activity of TiO₂
 nanoparticles" Appl. Catal. B: Environmental. 216 (2017) 80-87.
- 602 [39] G. Martra, E. Gianotti, S. Coluccia, The Application of UV Visible NIR Spectroscopy
- to Oxides, in: S.D. Jackson, J.S.J. Hargreaves (Eds.), Metal Oxide Catalysis, Wiley-VCH,
- 604 Weinheim, 2008, pp. 51-94.

- [40] R.I. Bickely, T. Gonzalez-Carreno, J.S. Lees, L. Palmisano, R.J.D. Tilley, A structural
 investigation of titanium dioxide photocatalysts, J Solid State Chem. 92 (1991) 178-190.
- 607 [41] O. Carp, C.L. Huisman, A. Reller, Photoinduced reactivity of titanium dioxide, Prog Solid
 608 State Ch. 32 (2004) 33-177.
- [42] N. Serpone, E. Pelizzetti, Photocatalysis Fundamentals and Applications, Wiley
 Interscience, New York, 1989.
- 611 [43] A. Marucco, F. Turci, L. O'Neill, H.J. Byrne, B. Fubini, I. Fenoglio, Hydroxyl density
- affects the interaction of fibrinogen with silica nanoparticles at physiological concentration, J
- Colloid Interf Sci. 419 (2014) 86-94.
- 614 [44] I. Fenoglio, G. Greco, S. Livraghi, B. Fubini, Non UV-induced radicals interactions at the
- surface of TiO₂ nanoparticles that may trigger toxic responses, Chem Eur J. 15 (2009) 4614-616 4621.
- [45] T. Daimon, T. Hirakawa, M. Kitazawa, J. Suetake, Y. Nosaka, Formation of singlet
 molecular oxygen associated with the formation of superoxide radicals in aqueous suspensions
 of TiO₂ photocatalysts, Appl Catal A. 340 (2008) 169–175.
- [46] A. Lipovsky, L. Levitski, Z. Tzitrinovich, A. Gedanken, and R. Lubart, The different
 behavior of rutile and anatase nanoparticles in forming oxy radicals upon illumination with
 visible light: an EPR study, Photochem Photobiol. 88 (2012) 14-20.
- [47] A. Marucco, E. Carella, I. Fenoglio, A comparative study on the efficacy of different
 probes to predict the photo-activity of nano-titanium dioxide toward biomolecules, RSC
 Advances. 5 (2015) 89559-89568.
- 626 [48] L. Attwood, D.M. Murphy, J.L. Edwards, T.A. Egerton, R.W. Harrison, An EPR study of
- 627 thermally and photochemically generated oxygen radicals on hydrated and dehydrated TiO2
- 628 surfaces, Res Chem Intermed. 29 (2003) 449-465.

- [49] C. Minero, V. Maurino, E. Pelizzetti, Mechanism of the photocatalytic transformation of
 organic compounds, in: V. Ramamurthy, K.S. Schanze, (Eds.), Semiconductor Photochemistry
 and Photophysics, Marcel Dekker, New York, 2003, pp. 211–229.
- [50] J. Cadet, T. Douki, J.L. Ravanat, Oxidatively generated damage to cellular DNA by UVB
- and UVA radiation, Photochem Photobiol. 91 (2015) 140-155.
- [51] Cadet J, T. Douki, J.L.Ravanat, Oxidatively generated damage to the guanine moiety of
- DNA: mechanistic aspects and formation in cells, Acc Chem Res. 41 (2008) 1075-1083.
- [52] J.P. Kehrer, L.O. Klotz, Free radicals and related reactive species as mediators of tissue
- 637 injury and disease: implications for health, Crit Rev Toxicol. 45 (2015) 765-798.

639	
640	
641	
642	
643	
644	
645	
646	
647	
648	
649	
650	
651	
	Supplementary information
652	Supplementally mormation
653	
654	
655	
656	
050	
657	
658	Indoor illumination: a possible pitfall in toxicological assessment of
659	photoactive nanomaterials.
660	
661	Arianna Marucco, Francesco Pellegrino, Simonetta Oliaro-Bosso, Valter Maurino, Gianmario
662	Martra, Ivana Fenoglio*
663	
664	
665	
666	
667	

Figure S1. Morphology of the samples. HRTEM micrographs of the four selected materials:
(A) UT001, (B) P25, (C) SX001 and (D) PC105. Scale bar in panels = 20 nm.



Figure S2. Adsorption spectra. DR UV-Vis spectra of: A) P25; B) UT001, C) PC105; D) SX001 in air. The Y axis is limited to 1.0, the limit of a correct application of the Kubelka-Munk function (Körtum, 1969).



Figure S3. XRD patterns of the samples



Figure S4. Reactivity of TiO₂ samples toward TEMPONE-H in different illumination conditions: A) filtered Hg/Xe lamp; B) reduced illumination C) dark. Panels on the left: generation of TEMPONE radicals by the TiO₂ samples (• UT001, • SX001, \Box P25, PC105, no powder) when in contact with a solution of TEMPONE-H (50 µM). The data are expressed as amount of radicals generated per unit surface area of the powder. Panels on the right: representative EPR spectra recorded after 40 minutes. (a) UT001, (b) SX001, (c) P25, (d) PC105, (e) no powder. The amount of radicals is proportional to the intensity of the signal.



Figure S5.Reactivity of TiO₂ samples toward sodium formate in different illumination conditions: A) filtered Hg/Xe lamp; B) reduced illumination C) dark. Panels on the left: generation of carboxylate radicals by the TiO₂ samples (● UT001, ○ SX001, □ P25, ▲ PC105, ■ no powder)in contact with a solution (0.005M PBS, pH 7.4, 88mM DMPO) of sodium formate (1M).The data are expressed as amount of radicals generated per unit surface area of the powder; Panels on the right: representative spectra recorded after 60 minutes (a) UT001, (b) SX001, (c) P25, (d) PC105, (e) no powder. The amount of radicals is proportional to the intensity of the signal.



Figure S6. Generation of singlet oxygen by TiO₂ samples. Panel on the left: amount of TEMPONE radicals generated by the TiO₂ samples (\bullet UT001, \circ SX001, \Box P25, \blacktriangle PC105, no powder) in a solution of 4-oxo-TMP (50mM) in phosphate buffer (pH 7.4, 0.01M) when illuminated with the filtered Hg/Xe lamp. Panel on the right: representative spectra recorded after 60 minutes (a) UT001, (b) SX001, (c) P25, (d) PC105, (e) no powder. The amount of radicals is proportional to the intensity of the signal.

733



734 735

736

737

Radical specie	Solvent	Hyperfine splitting constants
DMPO/CO ₂	0.005 mM phosphate buffer saline pH 7.4	a _H 15.4 G; a _N 18.5 G
TEMPONE	water	a _N 15.78 G
TEMPONE	0.01 mMphosphate buffer saline, pH 7.4	a _N 15.75 G

740 Table S1. Hyperfine splitting constants of the radical species detected