Title: Arbuscular mycorrhiza and plant succession on zinc smelter spoil heap in Katowice-Welnowiec

Author: : Ewa Gucwa-Przepióra, Katarzyna Turnau

Citation style: Gucwa-Przepióra Ewa, Turnau Katarzyna. (2001). Arbuscular mycorrhiza and plant succession on zinc smelter spoil heap in Katowice-Welnowiec. "Acta Societatis Botanicorum Poloniae" Vol. 70, no 2 (2001), s. 153-158, doi 10.5586/asbp.2001.020

ARBUSCULAR MYCORRHIZA AND PLANT SUCCESSION ON ZINC SMELTER SPOIL HEAP IN KATOWICE-WEŁNOWIEC

EWA GUCWA-PRZEPIÓRA[1], KATARZYNA TURNAU[2]

[1]Department of Plant Systematics, Silesian University
ul. Jagiellońska 28, 40-032 Katowice, Poland
e-mail: egucwa@us.edu.pl
[2]Insitute of Botany, Jagiellonian University
ul. Lubicz 46, 31-512 Kraków, Poland
e-mail: ubturnau@cyf-kr.edu.pl

(Received: February 9, 2001. Accepted: April 2, 2001)

ABSTRACT

Mycorrhizal status of plants colonising the zinc wastes in Katowice was surveyed. In total 69 vascular plant species (25 families) appearing on the investigated area have been noted. More than 60% of them were mycorrhizal. Non-mycorrhizal species, such as *Cardaminopsis arenosa* and *Silene vulgaris* dominated the early successional part of the zinc heap. *Tussilago farfara* was the only AM plant there, however, no arbuscules were developed at this stage. The number of AM species was increased on the 20 years old part of the zinc wastes and on the older 30-50 years old areas. AM plants constituted about 60% of the total number of species there. The frequency of particular AM species was the highest on the oldest part of the investigated area. The usefulness of the results for restoration practices was discussed.

KEY WORDS: arbuscular mycorrhiza, industrial wastes, heavy metals.

INTRODUCTION

Industrial wastes and strip-mines are a typical element of the Silesia region (South Poland). They cover over 3500 ha of land in the Katowice district and are often located close to large agglomerations. To overcome the problem of wind erosion, which creates a health hazard for local populations, the necessity to establish plant vegetation on the surface of the waste is an important goal. The substratum, however, is rather hostile for spontaneously appearing vegetation as it is often toxic to plant growth, low in mineral nutrient and lacking typical soil characteristics. The plant composition of the successional seres in this region is relatively well known (Rostański 1991, 1997a, b; Patrzałek and Rostański 1992; Tokarska-Guzik et al. 1991; Wika and Sendek 1993) documenting the dominance of wind dispersed, native, expansive species typical for anthropogenic environments (Faliński 1972; Rostański 1990). Due to the large area of the wastes the only possible remediation technique is on-site management, which would involve practices improving soil quality, introduction of properly selected plants and soil biota. In recent years substantial attention has been paid to mycorrhizal fungi which could be potentially used in revegetation of damaged ecosystems (Haselwandter 1997). While ectomycorrhizal fungi have already been successfully used in such cases, arbuscular fungi (AMF) still await the practical application despite the fact that this kind of mycorrhiza is the most important in plants which are useful for fast revegetation of the area. Although their contribution to increasing plant community production and structure, plant nutrition, soil structure development, diminishing water stress and increasing plant health status is well known, still the development of inoculation techniques, involving the selection of fungal strains adapted to toxic conditions under the given climatic and edaphic conditions is missing. The development of practices promoting growth and multiplication of AM fungi appearing spontaneously in such areas also requires the analysis of the mycorrhizal status of native plants colonising the wastes and the selection of those which would be the most effective in promoting fungal growth and stimulation. Although the mycorrhizal colonisation level can not serve as an indicator of the effectivity of the symbiosis, this parameter may be useful in monitoring the success of restoration or ecosystem stability (Lovera and Cuenca 1996). In addition, strongly colonised plant roots are one of the sources of the inoculum for other mycorrhizal plants, which will subsequently appear in the vicinity and thus contribute to the creation of more diverse vegetation system. Therefore, the aim of the present work was to study the mycorrhizal status of the plants spontaneously establishing during the successional stages of the revegetation of Katowice-Wełnowiec zinc wastes. This is considered as an important step in increasing the knowledge of mycorrhizal associations of this area and the start-point for further analysis of AM fungi diversity and selection of practically important fungal and plant species for restoration practices.

TABLE 1. Chemical properties of the zinc smelter spoil heap substrate in Katowice.

Part of heap	Below 20 years old	20 years old	30-40 years old	50 years old
pH	6.53	7.75	7.88	8.06
C organic [%]	9.92	8.73	6.24	3.14
N [%]	0.17	0.22	0.21	0.14
P_2O_5 [mg · 100 g^{-1} p.s.m.]	1.0	0.2	0.4	3.8
K_2O [mg · 100 g^{-1} p.s.m.]	4.4	7.75	6.3	9.5
Total metals [μg · g^{-1}]				
Zn	17410	17286	12901	7260
Pb	18666	17137	17996	11157
Cd	353	856	7023	173
Cu	1221	313	1549	1328
$Ca(NO_3)_2$ extractable metals [μg · g^{-1}]				
Zn	137.4	42.47	29.72	26.92
Pb	4.3	1.45	2.5	1.45
Cd	10.0	2.54	2.77	0.35
Cu	0.46	0.33	0.33	0.6

MATERIALS AND METHODS

Sampling site

The investigated zinc wastes which were studied in 1994 to 1997 are located in Wełnowiec, which is an industrial district of the city Katowice (Upper Silesia, southern Poland). The wastes occupy 25 ha (288-313 m above sea level). The slopes of the wastes were often very steep (40-45°) (Wojcierowska 1968). They are composed of floss from muffle furnaces, ash, waste materials from distillation and roasting furnaces and chamotte bricks. Presently, the area is abandoned.

Table 1 shows the physico-chemical soil characteristics of wastes of different age, reflecting the low content of organic matter, the low P and N content, and the high total heavy metal content.

Substratum sampling and analysis

Topsoil (0-15 cm depth) was sampled in May 1996. Five sub-samples were collected at random points from each waste of different age. Sub-samples were then bulked, sieved and air dried for soil extraction. Total and extractable (in 2N HNO_3 and 0.1 M $Ca(NO_3)_2$) elements in soils were determined with atomic absorption spectrometry (Ostrowska et al. 1991; Weissenhorn et al. 1995), N with the Kjeldahl method, organic matter with the Tiurin method, and available P and K with the Egner-Rhiem method (Lityński and Jurkowska 1982).

Plant sampling and preparation

The identification of plant species was carried out during the first vegetative season (from May to October). The plant material was deposited in the herbarium of the Department of Plant Systematics of the Silesian University (KTU). Latin names of plant species follow *Excursionsflora* (Rothmaler 1994). The percentage of plant cover and type of spatial distribution was estimated according to the commonly used Braun-Blanquet method (Braun-Blanquet 1964). Root samples of at least five plants from each sere of the zinc smelter spoil heaps, differing in age, were collected monthly during vegetative seasons 1994-1997. The roots were stained according to modified Phillips and Hayman's method (1970). The roots were washed in distilled water, cleared in 7% KOH at 60°C for 1 hour, rinsed in a few changes of water, acidified in 5% lactic acid at room temperature for 1-24 hours and stained in 0.05 trypan blue at 60°C for 0.5 hour. Stained roots were stored in lactoglycerol until they were used for slide preparation. Mycorrhizal colonisation and presence of AM structures (arbuscules, vesicles, coils and spores) were estimated according to the method described by Trouvelot et al. (1986).

RESULTS

AM status of the youngest wastes (under 20 years old)

The youngest zinc wastes were sparsely colonised by plants. Only ten species were noted there. The most common were non mycorrhizal *Cardaminopsis arenosa* and *Festuca ovina* (Table 2). *Tussilago farfara*, which was, however, much less common than the two mentioned species, was the only one where AM-like mycelia, coils and vesicles were found. No arbuscules were observed (Table 3).

TABLE 2. Non-mycorrhizal plant species from the zinc smelter spoil heap in Katowice
(parts of the heap are named as follows: 0 – up to 20 years old part, 1 – 20 years old part, 2 – 30-40 years old part, 3 – over 50 years old part).

Family	Species	Part of heap			
		0	1	2	3
Asteraceae	*Crepis biennis*		X	X	X
	Erigeron acer			X	X
	Eupatorium cannabinum		X	X	X
Boraginaceae	*Echium vulgare*		X	X	X
Brassicaceae	*Cardaminopsis arenosa*	X	X	X	
Caryophyllaceae	*Silene vulgaris*	X	X	X	X
Chenopodiaceae	*Atriplex patula*				X
	Chenopodium album			X	
Cyperaceae	*Carex hirta*				X
	Carex spicata			X	
Poaceae	*Agropyron repens*				X
	Calamagrostis epigeios	X	X	X	X
	Deschampsia caespitosa		X		
	Poa pratensis			X	
Polygonaceae	*Polygonum aviculare*			X	
	Reynoutria japonica				X
	Rumex acetosa	X	X	X	X
Resedaceae	*Reseda lutea*	X	X	X	X
Scrophulariaceae	*Verbascum densiflorum*			X	X
Solanaceae	*Solanum dulcamara*		X		
Urticaceae	*Urtica dioica*				X
Valerianaceae	*Valeriana officinalis*			X	

TABLE 3. AM presence and structures of plants from different parts of the wastes in Katowice
(parts of the heap are named as follows: 0 – up to 20 years old part, 1 – 20 years old part, 2 – 30-40 years old part, 3 – over 50 years old part).

Alphabetical list of plant species	Abundance and type of spatial distribution				AM presence	AM structures			
	0	1	2	3		A	C	V	S
Achillea millefolium		2.2		+.1	+	+	+	+	-
Agrostis capillaris		1.2			+/-	-	-	-	-
Agrostis stolonifera				2.2	+/-	+	-	+	-
Arrhenatherum elatius	1.2	+.2		1.2	+/-	+	+/-	+	-
Artemisia vulgaris				2.2	+	+	+	+	-
Campanula trachelium			r.1		+	-	+	+	+
Centaurea jacea		r.1			+	+	-	+	-
Chenopodium rubrum				+.2	+/-	-	+	+	-
Cirsium arvense			+.1	r.1	+/-	+	+/-	+/-	+/-
Cirsium vulgare			r.1	+.1	+	+	+	+	-
Convolvulus arvensis			+.1	+.1	+/-	+/-	+/-	+/-	-
Daucus carota		1.1	+.1	1.1	+/-	+	+/-	+	-
Deschampsia caespitosa		+.2			+	+	-	+	-
Epilobium angustifolium			+.1		+	+	-	+	+
Epilobium dodonaei			+.1		+	-	+	+	-
Erigeron acer			+.1	2.2	+/-	+	-	+	-
Euphorbia cyparissias		2.3			+	+	-	+	-
Festuca ovina		3.3	2.2	1.2	+/-	+	-	+	-
Hieracium lachenalii	2.3		+.1		+	+	+	+	-
Hieracium murorum		1.1			+	+	-	+	-
Hieracium piloselloides				1.1	+	+	-	+	-
Holcus lanatus			+.2		+	+	+	+	-
Leontodon autumnalis		2.2			+	+	-	+	-
Leontodon hispidus			+.1		+	+	-	+	-
Linaria vulgaris		2.2			+	+	+/-	+	+/-
Linum catharticum		r.1			+	-	+	-	-
Medicago lupulina				+.1	+	+	+	+	-
Melandrium album			1.2	1.2	+/-	+	-	+	-
Melilotus alba			+.1		+	+	-	+	+
Myricaria germanica			2.1		+/-	+	-	+	+
Oenothera biennis			r.1		+	+	+/-	+	-
Plantago lanceolata		1.2	2.2	2.2	+/-	+	+	+	+/-
Populus tremula		r.1	+.1	1.1	+/-	+	+/-	+	-
Reynoutria sachalinensis				1.3	+	+	-	+	-
Scrophularia nodosa		+.1	+.1		+	+	+	+	-
Solidago gigantea		1.2	1.2	2.3	+	+	+/-	+	-
Solidago virgaurea		+.1			+	+	+	+	-
Sonchus arvensis		r.1			+	+	+	+	+
Taraxacum officinale		1.1	+.1	1.1	+	+	+/-	+	-
Trifolium repens			1.3		+	+	+/-	+	-
Tripleurospermum inodorum				+.2	+	+	-	-	-
Tussilago farfara		2.2	+.2	+.2	+/-	+/-	+/-	+	-
Verbascum densiflorum	+.2		r.1	r.1	+	+	+/-	+	-
Vicia cracca			+.1	+.1	+/-	+	+/-	+	-
Viola tricolor		r.1		+.1	+	+	+	+	-

Abbreviations:
AM – arbuscular mycorrhiza; A – arbuscules; C – coils; V – vesicles; S – spores.
Each species' appearance is described by its abundance and type of spatial distribution in the patch according to the following rules: abundance: r – 1-3 individuals; + – the species covers up to 1% of the area; 1 – the species covers up to 5%; 2 – the species covers 5% to 25%; 3 – the species covers 25% to 50%; 4 – the species covers 50% to 75%; the species covers 75% to 100%; type of spatial distribution: 1 – single individuals, distributed randomly; 2 – small groups of individuals (rosettes and small tufts); 3 – bigger groups of individuals (tufts, small patches); 4 – big turfs, rug-like groups of individuals; 5 – corn-like pattern (very abundant occurrence).

20 years-old wastes

The vegetation developed on 20 years old wastes was much more abundant than on younger wastes. The relative cover of the ground reached 70%. It was dominated by nonmycorrhizal plants such as *Calamagrostis epigeios*, *Cardaminopsis arenosa* and *Reseda lutea* (Table 2). The number of mycorrhizal plant species increased to 60%; their population, however, was not abundant. The most common in this group were *Plantago lanceolata* and *Tussilago farfara*. Some plants were heavily colonised (Table 4). Arbuscules were observed in 58% of the observed plant species. In about 50% of these cases arbuscules were accompanied by coils (Table 3). The most abundant arbuscule formation was found in *Plantago lanceolata*, *Centaurea jacea*, *Daucus carota* and *Hieracium murorum* (Table 4).

30-40 years-old wastes

The vegetation cover at this stage was still about 70% but it was characterised by the highest richness of plant species. Among trees and shrubs the most common were *Betula pendula*, *Salix caprea* and *Myricaria germanica*, while among herbs *Cardaminopsis arenosa*, *Silene vulgaris* and *Festuca ovina* were still the most abundant. Mycorrhizal plants constituted 58% of

TABLE 4. Mycorrhizal colonization (M%) and arbuscules abundance (A%) of plants on different parts on the heap.
(values listed are means ± standard deviations; parts of the heap are named as follows: 0 – up to 20 years old part, 1 – 20 years old part, 2 – 30-40 years old part, 3 – over 50 years old part).

Species	M%	A%
Part 0		
Tussilago farfara	0.7 ± 1.7	0.0
Part 1		
Achillea millefolium	33.5 ± 9.4	23.9 ± 2.5
Arrhenatherum elatius	30.4 ± 8.9	8.0 ± 3.8
Centaurea jacea	87.8 ± 15.4	68.2 ± 16.2
Daucus carota	57.2 ± 6.4	56.0 ± 4.9
Deschampsia caespitosa	10.5 ± 6.3	6.5 ± 1.3
Euphorbia cyparissias	41.7 ± 15.4	18.3 ± 15.8
Festuca ovina	0.7 ± 1.5	0.0
Hieracium murorum	79.2 ± 22.6	59.4 ± 15.6
Leontodon autumnalis	16.6 ± 8.4	7.5 ± 3.6
Linaria vulgaris	66.9 ± 15.4	47.9 ± 10.0
Linum catharticum	1.5 ± 3.2	0.5 ± 3.6
Plantago lanceolata	20.2 ± 10.8	11.8 ± 1.8
Populus tremula	48.0 ± 15.4	0.0
Scrophularia nodosa	19.2 ± 10.4	9.6 ± 4.9
Solidago gigantea	43.5 ± 7.2	9.9 ± 1.5
Solidago virgaurea	28.1 ± 4.9	12.7 ± 7.2
Sonchus arvensis	67.3 ± 8.9	31.9 ± 4.9
Taraxacum officinale	11.6 ± 1.5	2.3 ± 0.2
Tussilago farfara	6.1 ± 10.4	2.8 ± 6.3
Viola tricolor	8.7 ± 2.7	2.1 ± 0.5
Part 2		
Campanula trachelium	0.7 ± 1.3	0.0
Cirsium arvense	2.6 ± 11.7	0.5 ± 7.7
Cirsium vulgare	64.1 ± 12.4	49.6 ± 13.4
Convolvulus arvensis	5.5 ± 2.5	2.2 ± 1.8
Daucus carota	0.7 ± 5.0	0.6 ± 1.5
Epilobium angustifolium	30.6 ± 6.3	15.4 ± 7.6
Epilobium dodonaei	82 ± 5.4	40.8 ± 12.6
Erigeron acer	0.6 ± 2.4	0.0
Festuca ovina	32.4 ± 13.8	20.1 ± 16.0
Hieracium lachenalii	10.5 ± 6.9	2.8 ± 5.1
Holcus lanatus	36.1 ± 7.6	23.2 ± 18.4
Leontodon hispidus	9.7 ± 6.4	3.9 ± 1.7
Melandrium album	2.6 ± 5.3	1.7 ± 3.1
Melilotus alba	46.4 ± 12.3	11.6 ± 7.6
Myricaria germanica	9.3 ± 18.1	7.7 ± 16.2
Oenothera biennis	43.6 ± 16.8	13.1 ± 5.9
Plantago lanceolata	55 ± 22.6	42 ± 8.4
Populus tremula	2.2 ± 5.1	1.6 ± 2.6
Scrophularia nodosa	58.4 ± 12.9	44.8 ± 7.2
Solidago gigantea	14.6 ± 10.1	3.7 ± 9.4
Taraxacum officinale	17.1 ± 9.8	5.9 ± 3.7
Trifolium repens	60.1 ± 19.8	44.2 ± 14.6
Tussilago farfara	37.4 ± 17.4	19.4 ± 8.9
Verbascum densiflorum	8.7 ± 13.2	2.1 ± 2.6
Part 3		
Achillea millefolium	34 ± 4.2	16.8 ± 3.3
Agrostis stolonifera	31.3 ± 4.5	16.3 ± 4.6
Arrhenatherum elatius	34.8 ± 5.4	4.0 ± 3.5
Arthemisia vulgaris	0.9 ± 13.9	0.1 ± 6.1
Chenopodium rubrum	0.7 ± 5.1	0.0
Cirsium arvense	2.6 ± 11.7	0.5 ± 7.7
Cirsium vulgare	10.7 ± 7.1	4.5 ± 3.4
Daucus carota	0.6 ± 5.3	0.0
Festuca ovina	2.6 ± 4.4	0.5 ± 1.7
Hieracium piloselloides	10.9 ± 4.2	4.1 ± 2.7
Medicago lupulina	62.0 ± 18.7	10.5 ± 4.7
Plantago lanceolata	41 ± 9.3	8.4 ± 5.0
Populus tremula	16.9 ± 8.4	4.1 ± 5.6
Reynoutria sachalinensis	0.9 ± 3.2	0.1 ± 1.8
Solidago gigantea	19.6 ± 7.4	13.5 ± 3.7
Taraxacum officinale	7.7 ± 4.2	4.1 ± 2.0
Tripleurospermum inodorum	21.8 ± 6.3	10.2 ± 5.2
Tussilago farfara	7.9 ± 3.4	0.0
Verbascum densiflorum	0.5 ± 2.8	0.3 ± 3.7
Vicia cracca	29.8 ± 11.2	13.4 ± 4.6
Viola tricolor	13.1 ± 5.6	4.9 ± 2.8

plant species. However, only some of them grew in abundant populations e.g. *Myricaria germanica* (Table 3). Nonmycorrhizal plants (42% of the number of species) such as *Silene vulgaris*, *Calamagrostis epigeios*, *Rumex acetosa* and *Cardaminopsis arenosa* were still very frequent in this part (Table 2). Roots of *Plantago lanceolata*, *Scrophularia nodosa* and *Cirsium vulgare* were heavily colonised, and also the arbuscule richness was high (Table 4). The arbuscule richness was lower than on the 20 years old stage and did not exceed 50% (Table 4). The number of species with low abundance of arbuscules was twice increased.

50 years-old zinc wastes

The vegetation of this stage differed from the previous ones in richness and ground cover (up to 85%). It was dominated by abundant populations of *Solidago gigantea*, *Plantago lanceolata*, *Arthemisia vulgaris*, *Achillea millefolium* and *Daucus carota* forming big tufts and clusters (Table 3). Mycorrhizal species made up to 59% of the total number of plants, as in the case of the 30-40 years-old wastes. At the same time, however, these species dominated the vegetation. Arbuscules were noted in 55% of species, however, plants with low arbuscule richness (up to 10%) dominated. The value of this parameter did not exceed 20% (Table 4). Some nonmycorrhizal species, which were common in youngest stages did not appear in this part (e.g. *Cardaminopsis arenosa*) (Table 2).

DISCUSSION

Despite the large areas taken by industrial wastes in Central Europe the investigations on the mycorrhiza of plants colonising such places are comparatively rare. So far such data are available for Poland from wastes of electricity power plant (Turnau 1987) soda factory spoil mound (Pawłowska 1991), calamine wastes (Pawłowska et al. 1996; Turnau et al. 1996) and zinc wastes (Turnau 1998; Turnau et al. 2000). Although heavy metal containing wastes have already been included in many investigations the present paper is the first one to give the characteristic of the successional stages occurring during 50 years from the moment of the waste deposition. The estimation of the age was possible on the basis of the documentation of the area. Moreover, chemical characterisation of the substratum, carried out within this research provided additional information which could be important for the comparison of data and could allow for avoidance of wrong conclusions concerning the substrata which could have originated as a result of different technologies used. In the present research 69 plant species appearing on the zinc wastes were analysed. 47 were found to form arbuscular mycorrhiza including 15 that were facultative hosts for AM fungi. According to the available literature the mycorrhizal status of 7 species of vascular plants found in this area have been studied for the first time. No mycorrhizal colonisation has been found in the case of *Carex spicata* and *Reynoutria japonica*, while *Myricaria germanica*, *Epilobium dodonaei*, *Hieracium piloselloides*, *Verbascum densiflorum*, *Solidago gigantea*, evidently form arbuscular mycorrhizas, including the formation of fully developed arbuscules.

The zinc waste of Wełnowiec is an example of an area on which, due to high heavy metal content, the infertility of the substratum and the lack of the appropriate soil structure the spontaneous revegetation is a very long process. High heavy metal content seems to be the main problem on the youngest wastes (under 20 years old). This part of the waste was characterised by the highest content of extractable metals such as Zn, Pb and Cd. This was correlated with the lack of arbuscules within *Tussillago farfara* roots, which in the later stages of succession developed arbuscular mycorrhiza. The mycelium forming coils and vesicles was the only sign of AMF presence in the youngest stage of waste revegetation. The following two successional stages were characterised by much more abundant arbuscule formation while at the last stage this parameter was strongly decreased. This might be caused by the much higher availability of Cu in these wastes; however, other reasons could not be excluded.

The above described plant succession on heavy metal containing wastes generally followed the classical model proposed by Janos (1980) for disturbed habitats with the early dominance of nonmycorrhizal plants, later replaced with facultative and finally obligatory mycorrhizal species. In the case of the Wełnowiec wastes the first stage was dominated by nonmycorrhizal plants which are mostly known as heavy metal accumulating plants. *Silene vulgaris*, *Cardaminopsis arenosa* and *Agrostis capillaris* are typical examples of this plant group. In addition, this stage was much longer than in the so far published cases (Allen and Allen 1980) what was probably due to especially severe conditions resulting in the delay of AM fungi establishment. The role of AM fungi in revegetation of mine wastes is comparatively well known (Daft and Hacskaylo 1976; Daft and Nickolson 1974; Allen 1989a, b; Miller and Jastrow 1992; Allen and Allen 1990), however, so far no restoration practices involving this group of fungi were carried out in Poland. Basing on experiences obtained in other regions the rate of restoration may be increased by the manipulation of the mycorrhizal fungal population or the inoculation techniques (Reeves et al. 1979; Janos 1980). The usefulness of indigenous AMF used as inoculants to reclaim mine spoils (Khan 1981; Stahl et al. 1988) and oil polluted soil (Call and McKell 1982; Cabello 1995, 1997, 1999) has been shown by several authors. The number of AMF propagules depends on factors such as soil nutritional status, host plant, AMF propagule density, effectiveness of AMF species and competition between them and other soil microorganisms (Cabello 1999). Fungi isolated from polluted areas were more effective than those originating from unpolluted sites, suggesting the adaptation of fungi to persistent toxicants in soil. The AM fungi from the Wełnowiec waste could be used in future for production of inoculum well adapted for edaphic and climatic conditions and by this better adapted for restoration practices in the particular area. Before the introduction of the inoculum in the first successional stages it seems to be necessary to use appropriate techniques lowering the toxicity of the substratum and increasing the nutritional status. Optimalisation of both should be carried out on the basis of experiments where the AM fungal activity would be monitored using the spore germination test as described by Weissenhorn et al. (1993) or by checking mycorrhiza development of selected plants. In the part of the waste which is already colonised by mycorrhizal fungi the restoration techniques should enhance mycorrhiza development, for example by applying appropriate levels and kinds of organic amendments which again should be optimalised by experimental trials. Another aim of phytoremediation of the waste is the selection of appropriate plant species. The so far developed techniques often involve nonmycorrhizal plant species. The growth of plants such as *Silene vulgaris* and *Thlaspi caerulescens* had, however, a more negative effect on the number of AM spores in the substratum than while mycorrhizal *Zea mays* was present (Pawłowska et al. 2000). In the case of the Wełnowiec wastes the native plant species could be selected on the basis of above presented data concerning their mycorrhizal status.

ACKNOWLEDGMENTS

Special thanks are due to Dr Ryszard Ciepał (Department of Ecology, Silesian University, Katowice) and Dr Barbara Godzik (Polish Academy of Sciences, Kraków) for the help in estimation of heavy metal levels and it availability, to eng. Maciej Terakowski (Institute for Ecology of Industrial Areas, Katowice) for the help in soil analysis and to Dr Anna Jurkiewicz for her linguistic comments during the preparation of the manuscript.

The research was supported by the Committee for Scientific Research in Poland, grant no. PB 0281/PO4/95/08.

LITERATURE CITED

ALLEN E.B. 1989a. The restoration of disturbed arid landscapes with special reference to mycorrhizal fungi. J. Arid Environm. 17: 279-286.

ALLEN M.F. 1989b. Mycorrhizae and rehabilitation of disturbed arid soils: Processes and practices. Arid Soil Res. 3: 229-241.

ALLEN E.B., ALLEN M.F. 1980. Natural re-establishment of vesicular-arbuscular mycorrhizae following stripmine reclamation in Wyoming. J. Appl. Ecol. 17: 139-147.

ALLEN E.B., ALLEN M.F. 1990. The mediation of competition by mycorrhizal fungi in successional and patchy environments. In: Grace J.R., Tilman G.D. (eds) Perspectives in plant competition. Academic Press, New York, pp. 367-389.

BRAUN-BLANQUET J. 1964. Pflanzensozologie, Grundzüge der Vegetationskunde. 3 Aufl. Springer Verl., Wien–New York.

CABELLO M.N. 1995. Effect of hydrocarbon pollution on vesicular-arbuscular mycorrhizal fungi (VAM) Bol. Micol. (Chile) 10: 77-83.

CABELLO M.N. 1997. Hydrocarbon pollution: its effect on native arbuscular mycorrhizal fungi (AMF). FEMS Microbiol. Ecol. 22: 233-236.

CABELLO M.N. 1999. Effectiveness of indigenous arbuscular mycorrhizal fungi (AMF) isolated from hydrocarbon polluted soils. J. Basic Microbiol. 39(2): 89-95.

CALL C.A., MCKELL C.M. 1982. Vesicular-arbuscular mycorrhizae – a natural revegetation strategy for disposed processed oil shale. Reclam. Revegn. Res. 1: 337-347.

DAFT M.J., HACSKAYLO E. 1976. Arbuscular mycorrhizas in anthracite and bituminous coal wastes of Pennsylvania. J. Appl. Ecol. 13: 523-531.

DAFT M.J., NICKOLSON T.H. 1974. Arbuscular mycorrhizas in plants colonizing coal wastes in Scotland. New Phytol. 73: 1129--1137.

FALIŃSKI J.B. 1972. Synantropizacja szaty roślinnej – próba określenia istoty procesu i głównych kierunków badań. Phytocenosis 1(3): 157-169.

HASELWANDTER K. 1997. Soil micro-organisms, mycorrhiza, and restoration ecology. In: Urbańska K., Webb N.R., Edwards P.J. (eds) Restoration ecology and sustainable development. Cambridge Univ. Press, Cambridge, pp. 65-80.

JANOS D.P. 1980. Mycorrhizae influence tropical succession. Biotropica 12: 56-64.

KHAN A.G. 1981. Growth responses of endomycorrhizal onions in unsterilized coal wasts. New Phytol. 87: 363-370.

LOVERA M., CUENCA G. 1996. Arbuscular mycorrhizal infection in Cyperaceae and Gramineae from natural, disturbed and restored savannas in La Gran Sabana, Venezuela. Mycorrhiza 6: 111-118.

LITYŃSKI T., JURKOWSKA H. 1982. Żyzność gleby a odżywianie się roślin. PWN, Warszawa.
MILLER R.M., JASTROW J.D. 1992. The application of VA mycorrhizae to ecosystem restoration and reclamation. In: Allen M.F. (ed.) Mycorrhizal functioning: an integrative plant-fungal process. Chapman and Hall, New York, pp. 438-467.
OSTROWSKA A., GAWLIŃSKI S., SZCZUBIAŁKA P. 1991. Metody analizy i oceny właściwości gleb i roślin. Katalog. Instytut Ochrony Środowiska, Warszawa.
PATRZAŁEK A., ROSTAŃSKI A. 1992. Procesy glebotwórcze i zmiany roślinności na skarpie rekultywowanego biologicznie zwałowiska odpadów po kopalnictwie węgla kamiennego. Arch. Ochr. Środ. 3-4: 157-168.
PAWŁOWSKA T. 1991. Plant mycorrhizae in the sedimentation tanks of the Cracow Soda Factory. Zesz. Nauk. Uniw. Jagiellon., Prace Bot. 22: 163-170.
PAWŁOWSKA T.E., CHANEY R.L., CHIN L., CHARVAT I. 2000. Effects of metal phytoextraction practices on the indigenous community of arbuscular mycorrhizal fungi at a metal-contaminated landfill. Appl. Environm. Microbiol. 66 (6): 2526-2530.
PAWŁOWSKA T.E., BŁASZKOWSKI J., RÜHLING A. 1996. The mycorrhizal status of plants colonizing a calamine spoil mound in southern Poland. Mycorrhiza 6: 499-505.
PHILLIPS J.M., HAYMAN D.S. 1970. Improved procedures for clearing roots and staining parasitic and vesicular-arbuscular mycorrhizal fungi for rapid assesment of infection. Trans. Brit. Mycol. Soc. 55: 158-160.
REEVES F.B., WAGNER D.W., MOORMAN T., KIEL J. 1979. The role of endomycorrhizae in revegetation practises in the semi-arid west. I. A comparison of incidence of mycorrhizae in severly disturbed vs. natural environments. Am. J. Bot. 66: 1-13.
ROSTAŃSKI A. 1991. Spontaniczna sukcesja roślinności na wybranych zwałach poprzemysłowych w województwie katowickim. Kształtowanie środowiska geograficznego i ochrona przyrody na obszarach uprzemysłowionych 3: 35-38.
ROSTAŃSKI A. 1997a. Flora spontaniczna hałd Górnego Śląska. Arch. Ochr. Środ. 23 (3-4): 159-165.
ROSTAŃSKI A. 1997b. Zawartość metali ciężkich w glebie i roślinach z otoczenia niektórych emitorów zanieczyszczeń na Górnym Śląsku. Arch. Ochr. Środ. 23 (3-4): 181-189.
ROSTAŃSKI K. 1990. Skutki antropopresji we florze naczyniowej regionu uprzemysłowionego na przykładzie Górnośląskiego Okręgu Przemysłowego i terenów sąsiednich. In: Godzik S. (ed) Zagrożenie i stan środowiska przyrodniczego rejonu śląsko-dąbrowskiego 62: 58-69.
ROTHMALER W. 1994. Excursionsflora von Deutschland. 4. Gefässpflanzen: Kritischer Band. G. Fischer Verl., Jena, Stuttgart.
STAHL P.D., WILLIAMS S.E., CHRISTENSEN M. 1988.Efficacy of native vesicular-arbuscular mycorrhizal fungi after severe soil disturbance. New Phytol. 110: 347-354
TOKARSKA-GUZIK B., ROSTAŃSKI A., KLOTZ S. 1991. Roślinność hałdy pocynkowej w Katowicach-Wełnowcu. Acta Biol. Siles. 19(36): 94-102.
TROUVELOT A., KOUGH J.L., GIANINAZZI-PEARSON V. 1986. Mesure du taux de mycorhization VA d'un systéme radiculaire. Recherche de méthodes d'estimationayant une signification fonctionelle. Mycorrhizae: physiology and genetics. ESM 1: 217-221.
TURNAU K. 1987. Changes in mycoflora during revegetation of fly-ash heap at Skawina (southern Poland). Zesz. Nauk. Uniw. Jagiellon., Prace Bot. 15: 159-163.
TURNAU K. 1998. Heavy metal uptake and arbuscular mycorrhiza development of *Euphorbia cyparissias* on zinc wastes in South Poland. Acta Soc. Bot. Pol. 67: 105-113.
TURNAU K. KOTTKE I. DEXHEIMER J. 1996. Toxic element filtering in *Rhizopogon roseolus/Pinus sylvestris* mycorrhizas collected from calamine dumps. Mycol. Res. 100 (1): 16-22.
TURNAU K., RYSZKA P., VAN TUINEN D., GIANINAZZI-PEARSON V. 2000. Identification of arbuscular mycorrhizal fungi in soils and roots of plants colonizing zinc wastes in Southern Poland. Mycorrhiza. (in press)
WEISSENHORN I., LEYVAL C., BERHELIN J. 1993. Cd-tolerant arbuscular mycorrhizal (AM) fungi from heavy metal polluted soil. Plant & Soil 157: 247-256.
WEISSENHORN I., MENCH M., LEYVAL C. 1995. Bioavibility of heavy metals and abundance of arbuscular mycorrhiza in a sewage sludge amended sandy soil. Soil Biol. Biochem. 27: 287-296.
WIKA S., SENDEK A. 1993. Sukcesja swoistej roślinności na hałdzie hutniczej w Siemianowicach Śląskich. Kształtowanie środowiska geograficznego i ochrona przyrody na obszarach uprzemysłowionych i zurbanizowanych 9: 23-30.
WOJCIEROWSKA M. 1968. Projekt wstępny Nr 97-544 rekultywacji i zagospodarowania terenów zwałowisk (M.S.). Katowice, Huta „Silesia".

MIKORYZA ARBUSKULARNA I SUKCESJA ROŚLINNA NA HAŁDZIE POCYNKOWEJ W KATOWICACH-WEŁNOWCU

STRESZCZENIE

Zbadano status mikoryzowy roślin hałdy pocynkowej w Katowicach-Wełnowcu. Odnotowano 69 gatunków (z 25 rodzin). Ponad 60% z nich tworzyło mikoryzę. Gatunki niemikoryzowe, jak *Cardaminopsis arenosa* i *Silenene inflata*, dominowały na młodszych częściach hałdy. *Tussilago farfara* był tam jedynym gatunkiem mikoryzowym, jednak nie stwierdzono u niego arbuskul. Liczba gatunków mikoryzowych wzrosła na starszych częściach. Rośliny mikoryzowe stanowiły tam około 60% wszystkich gatunków. Częstość występowania gatunków mikoryzowych była najwyższa na najstarszej części hałdy. Przedyskutowano możliwość wykorzystania otrzymanych wyników w rekultywacji tego typu terenów.

SŁOWA KLUCZOWE: mikoryza arbuskularna, hałdy przemysłowe, metale ciężkie.