

1 **Assessing landscape dust emission potential using combined ground-based**
2 **measurements and remote sensing data**

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15

16 **Key points:**

- 17
- 18 • The combination of remote sensing and ground-based measurement is potent for
19 studying dust emission potential across spatial scales
 - 20 • Results demonstrate substantial variability of emission at each scale of analysis
(individual erosional surface, landform, and landscape)
 - 21 • A Boosted Regression Tree model determines the relative influence of specific
22 variables controlling surface erodibility

23

24 **Abstract**

25 Modeled estimates of aeolian dust emission can vary by an order of magnitude due to the
26 spatiotemporal heterogeneity of emissions. To better constrain location and magnitude of
27 emissions, a surface erodibility factor is typically employed in models. Several landscape-scale
28 schemes representing surface dust-emission potential for use in models have recently been
29 proposed, but validation of such schemes has only been attempted indirectly with medium-
30 resolution remote sensing of mineral aerosol loadings and high-resolution land-surface
31 mapping. In this study, we used dust-emission source points identified in Namibia with Landsat
32 imagery together with field-based dust-emission measurements using a Portable In-situ Wind
33 Erosion Laboratory (PI-SWERL) wind tunnel to assess the performance of schemes aiming to
34 represent erodibility in global dust-cycle modeling. From analyses of the surface and samples
35 taken at the time of wind tunnel testing, a Boosted Regression Tree analysis identified the
36 significant factors controlling erodibility based on PI-SWERL dust flux measurements and
37 various surface characteristics, such as soil moisture, particle size, crusting degree and
38 mineralogy. Despite recent attention to improving the characterisation of surface dust-emission
39 potential, our assessment indicates a high level of variability in the measured fluxes within
40 similar geomorphologic classes. This variability poses challenges to dust modelling attempts
41 based on geomorphology and/or spectral-defined classes. Our approach using high-resolution
42 identification of dust sources to guide ground-based testing of emissivity offers a valuable
43 means to help constrain and validate dust-emission schemes. Detailed determination of the
44 relative strength of factors controlling emission can provide further improvement to regional
45 and global dust-cycle modeling.

46 **Plain Language Summary**

47 Atmospheric mineral dust plays an important role in Earth system processes, influencing
48 climate, providing nutrients to ecosystems and affecting human health. The effect that
49 atmospheric dust has on the climate and environment requires accurate modeling of emission
50 at source, transport through the atmosphere, and deposition. To enable regional to global
51 modeling of the dust cycle, therefore, requires realistic representation of where and when dust
52 emission takes place. However, the highly variable nature of dust emission has resulted in
53 modeling attempts producing disparate results. This research used Landsat remote sensing data
54 in Namibia to identify sources of dust emission at high-resolution, followed by ground-based
55 testing using a portable wind tunnel to assess surface classification schemes intended to

56 represent the surface in dust-emission models. Despite the proposed schemes offering valuable
57 approaches for characterization of the land surface for modeling, globally applicable
58 representation of dust emission is still hampered by the variability of small-scale emissions. At
59 the sub-landform level of our analysis, the heterogeneous nature of dust emission results from
60 the highly variable nature of the surfaces. Our analysis identified several factors controlling the
61 potential for surfaces to emit dust that can be used as inputs to improve dust modeling.

62 **1 Introduction**

63 Wind-driven processes of sediment transport are important in the Earth system and
64 consequently have been the focus of many modeling attempts (Ravi et al., 2011; Shao et al.,
65 2011). The dynamics of mineral dust emission are fundamentally controlled by a combination
66 of the power of the wind to erode (erosivity) and the resistance of an emitting surface to erosion
67 (erodibility) (Webb & Strong, 2011). Interactions between erosive and resisting forces are
68 complex and result in dust emission being spatially and temporally highly heterogeneous (e.g.
69 Bryant et al., 2007; Gillette, 1999; Gillies, 2013; Mahowald et al., 2003; Taramelli et al., 2013).
70 Improvements in dust-emission modeling remains an important contemporary research goal
71 since existing models have a limited capacity to accurately account for the spatiotemporal
72 variability of dust emission within dust sources (Haustein et al., 2015; Parajuli et al., 2014;
73 Shao et al., 2011).

74 Modeling of dust emission must account for factors that affect the threshold friction velocity
75 (u_{*t}), and as a result, the variable erodibility of the surface (Marticorena & Bergametti, 1995,
76 Shao et al., 1996). Some of the major drivers influencing the variability of the surface erosion
77 thresholds include soil moisture (influenced by relative humidity), particle size, degree of
78 crusting (including physical, saline and biological soil crusts), and mineralogy of surface
79 sediments (e.g. Belnap & Gillette, 1998; Buck et al., 2011; Cornelis et al., 2004; Gillette et al.,
80 1982; King et al., 2011; Marticorena & Bergametti, 1995; McKenna Neuman & Nickling,
81 1989; McKenna Neuman & Maxwell, 2002; Munkhtsetseg et al., 2016; Sweeney et al., 2016),
82 as well as surface roughness (characterized by the aerodynamic roughness length, z_o) (Raupach
83 et al. 1993), with vegetative and topographic (micro to macro) roughness having significant
84 influences (e.g. Gillies et al., 2006; Okin & Gillette, 2001; Sankey et al., 2010). Incorporating
85 the influence of these surface characteristics into soil erodibility and dust emission predictions
86 is one of the biggest challenges for dust simulation, especially given that global data sets of

87 these input variables are not always available, or are not at a spatial scale appropriate for model
88 input.

89 A surface erodibility factor is typically used in dust models to constrain the observed spatial
90 heterogeneity of emissions (Zender et al., 2003). Several dust-emission mapping schemes at
91 the landscape scale have attempted to account for erodibility as a regulator of emission
92 potential for use in dust models (e.g. Ashpole & Washington, 2013; Baddock et al., 2016;
93 Bullard et al., 2011; Parajuli et al., 2014; Parajuli & Zender, 2017). The erodibility factor has
94 typically been based on various physical assumptions of the influence of geomorphology,
95 topography and hydrology on dust emission (Ginoux et al., 2001; Zender et al., 2003).
96 Alternatively, empirical approaches based on satellite-derived data, including surface
97 reflectance (e.g. Grini et al., 2005) have also been formulated. Bullard et al. (2011) and Parajuli
98 et al. (2014) presented high resolution land-surface classifications based on the potential
99 emissivity of specific geomorphic types and land covers. A recent global characterization of
100 dust-emission potential by Parajuli & Zender (2017), the Sediment Supply Map (SSM),
101 combines drainage area (a proxy for long-term hydrologic transport and deposition of
102 sediment) with empirically-derived surface reflectance from the Moderate Resolution Imaging
103 Spectroradiometer (MODIS) blue channel. The combination of these datasets encapsulates two
104 important aspects of sediment supply, namely the accumulation of fine sediments in basins as
105 a supply of dust-sized material, and the reflectance of different land surface types based on
106 their surface sediment supply potential (Parajuli & Zender, 2017). The SSM is a landscape-
107 scale (~500 m) erodibility map that provides numerical estimates of dust emission potential
108 for use in global dust-cycle models.

109 While such classifications are produced at a high spatial resolution relative to current dust
110 modeling approaches (Shi et al., 2016; Parajuli & Zender, 2017), it is recognised that a range
111 of influences affecting dust emission operate at scales below the landscape scale. As such, the
112 scale at which dust-emission processes are investigated has a marked influence on the spatial
113 representation of emission variability. Webb & Strong (2011) highlight this by proposing that
114 wind erosion drivers can be understood at a range of scales, with different influences apparent
115 at successive scales of analysis: grain to surface ($<10^0$ m), landform ($\sim 10^1 - 10^2$ m), landscape
116 ($\sim 10^3$ m), and regional to global scales ($>10^4$ m). Recent landscape-scale dust-emission mapping
117 schemes have not yet been assessed rigorously by ground-truthing and uncertainty remains
118 regarding how well these surface classifications account for the potential variability in emission
119 known to exist at the landform and sub-landform scales (Sweeney et al., 2011).

120 Our understanding of dust-emission processes has been greatly enhanced by studies that have
121 identified dust sources on global, regional and landscape scales through various remote sensing
122 approaches primarily using the Total Ozone Mapping (TOMS) and more recently the MODIS
123 sensors (e.g. Baddock et al., 2016; Bullard et al., 2008, Ginoux et al., 2012; Huang et al., 2007,
124 Lee et al, 2012, O’Loingsigh et al., 2015; Prospero et al., 2002; Schepanski et al., 2007, 2012;
125 Vickery et al., 2013; Washington et al., 2003). However, a fuller appreciation of the smaller-
126 scale controls contributing to the variability in dust emission also depends on the improved
127 characterisation of dust sources at a sub-landform scale. Ground-based studies are crucial,
128 because the sub-landform variability of emission from dust producing surfaces has proven
129 difficult to investigate using other means. However, using ground-based measurements to
130 validate predictions of dust emission potential (such as that provided by the SSM) remains a
131 challenge, because of the disconnect between process studies and flux measurements,
132 necessarily performed at a landform to sub-landform scale, *versus* the regional or global focus
133 taken by modeling studies. This is partly due to the limitation posed by a relatively coarse
134 spatial resolution in remote sensing together with a lack of dedicated field studies quantifying
135 sub-landform variability (Haustein et al., 2015). Small-scale studies allow quantification of
136 dust emission from specific landforms and the combination of surfaces within these landforms.
137 The advantage of a high-resolution approach to dust source-point identification has recently
138 been demonstrated by von Holdt et al. (2017) who used Landsat imagery covering a 25-year
139 period to identify the landform-scale dust sources in the Namib Desert of southern Africa.

140 The increased spatial resolution of Landsat (15-30 m) compared to other remote sensing data
141 used to date (e.g. MODIS 250-1000 m) has improved accuracy for dust source-point
142 identification, allowing the study of dust emission at landform scales and guiding field
143 measurement at the sub-landform scale (von Holdt et al., 2017). The spatial variability of dust
144 emission at sub-landform scale has been investigated by several studies using a PI-SWERL
145 wind tunnel (Etyemezian et al., 2007) to measure the dust emission potential of surfaces from
146 a variety of landforms found in desert regions (e.g. Bacon et al., 2011; King et al., 2011;
147 Sweeney et al., 2011; 2016). The small size and portable nature of this instrument allows for
148 replicate testing of multiple surfaces in locations that would not be accessible by conventional,
149 larger footprint wind tunnels. Furthermore, given the size of the PI-SWERL (0.57 m diameter),
150 its measurements are at a spatial scale corresponding to the grain and surface scale controls on
151 dust emission ($<10^0$ m) (Webb and Strong, 2011). Using a Landsat analysis to guide in situ
152 measurements for quantifying surface- to landscape-scale variability of dust emission (von

153 Holdt et al., 2017), offers a means for testing dust-emission schemes and improving how
154 surface erodibility is characterized in dust modeling.

155 This study aims to use a portable wind tunnel to estimate relative emissivity from different
156 land surface types, doing so within the context of recently proposed methods for classifying
157 surface emission potential for dust modeling efforts. Assessment of measured dust fluxes
158 from classified surfaces is used to contribute a novel test of these new schemes and more
159 broadly, inform regional and global dust models. For flux measurements, field-based
160 emission sampling with a PI-SWERL was guided by using high a resolution, Landsat-
161 derived, dust source point inventory created for the Namib Desert (von Holdt et al., 2017).
162 This approach allows assessment of emission variability across a range of spatial scales by
163 combining PI-SWERL point measurements with landform classification. The secondary
164 objective was to examine the emission measurements and a range of surface properties (soil
165 moisture, degree of crusting, particle size and mineralogy) using a Boosted Regression Tree
166 (BRT) analysis to determine the most significant erodibility factors for the dust source points.

167 **2 Regional setting and field sites**

168 The Namib Desert is one of the major southern African dust sources (Vickery et al., 2013; von
169 Holdt et al., 2017), and is appreciable at the hemispheric scale (Ginoux et al., 2012). This region
170 comprises several desert landforms, including 12 westward-flowing ephemeral rivers,
171 numerous small inland playas and large coastal sabkhas, sand deposits which include sand
172 sheets and sand dunes; and extensive areas of stony desert comprising gravel stone pavements
173 dissected by non-fluvial ephemeral drainage channels (Bullard et al., 2011; Jacobsen et al.,
174 1995; Goudie & Viles, 2015). Dust emission from the Namib Desert has been mostly associated
175 with the terminal stages of the dry river valleys and coastal sabkhas and inland playas (Dansie
176 et al., 2018; Eckardt & Kuring, 2005; Vickery & Eckardt, 2013; von Holdt et al., 2017). The
177 Kuiseb, Huab and Omaruru rivers were identified as the most emissive river systems based on
178 MODIS true colour imagery analysis from 2005 to 2015, whereas Conception Bay and the
179 Ugab Pans were the most emissive sabkhas (von Holdt et al., 2017) (Figure 1 a). The present
180 study uses PI-SWERL measurements to assess dust emission potential predicted from
181 classification schemes applied to the three most emissive catchments determined by von Holdt
182 et al. (2017), in addition to the Ugab sabkha system (marked U in Figure 1a). The PI-SWERL
183 measurements from von Holdt et al. (2017) are a subset of the data used in the present paper
184 (40% of the total data set, Table S1), and while von Holdt et al. (2017) examined the river

185 systems on a case-by-case basis, the current investigation examines dust emission across
186 multiple scales (erosional surface, landform, and landscape) for the Namib Desert study area.

187 **3 Methods**

188 3.1 Geomorphology and dust emission scheme mapping

189 Geomorphological units were mapped in the study area following the land-surface
190 classification based on geomorphology used by Bullard et al. (2011) in their Preferential Dust
191 Scheme (PDS) (see also Baddock et al., 2016). The PDS classes included lake systems,
192 including dry and ephemeral lakes (playa and sabkha pans), alluvial systems (high- and low-
193 relief), stony systems (including stone pavements intersected by ephemeral drainage channels),
194 aeolian systems (sand sheets and dunes), loess and low emission surfaces, such as bedrock.
195 The Namib loess deposits consist predominantly of fluvially reworked loess in the ephemeral
196 river valleys (Eitel et al., 2001) and were mapped as part of alluvial systems as they are not
197 distinguishable at the scale of mapping used in the present study. The study area mapped
198 consisted of the Landsat tiles analysed by von Holdt et al. (2017) (Figure 1 a) and used a
199 combination of remote sensing data, 1:250 000 geological maps from the Geological Survey
200 of the Ministry of Mines and Energy of Namibia and field observations. The remote sensing
201 data included Google Earth images, Landsat 8 false colour imagery (bands 7,5,3) and the
202 Shuttle Radar Topography Mission (SRTM) 30-m digital elevation model to distinguish
203 between low- and high-relief, as well as degree of incision of alluvial systems. The 2289 dust
204 source points identified with the aid of Landsat imagery between 1990 and 2016 by von Holdt
205 et al. (2017) were classified according to the PDS land-surface classes at a landscape scale.
206 Mapping was done in QGIS v 2.18.12 (QGIS development team, 2016).

207 The Land Surface Map (LSM) and Sediment Supply Map (SSM) were made available as rasters
208 by Parajuli & Zender (2017). The LSM (Parajuli et al., 2014) was originally developed by
209 mapping the Middle East and North Africa region according to 12 spectral land cover classes
210 with high-resolution Google Earth Pro images and polygons created as training samples for a
211 global supervised classification which used the maximum likelihood method in ArcGIS, as
212 applied to the global Blue Marble (MODIS RGB) image mosaic. To enable better comparison
213 of the PDS and LSM outputs, the Parajuli et al. (2014) spectral land cover classes were
214 reclassified according to the geomorphology based PDS land-surface classes (Bullard et al.,
215 2011). The LSM is used for a qualitative and quantitative comparison with the SSM produced

216 globally by Parajuli & Zender (2017). The original LSM land cover classes used in Parajuli et
217 al. (2014) are included in the Supporting Information Figure S1.

218 The SSM is derived through a combination of the upstream catchment area and the surface
219 reflectance captured in the blue band (459-479 nm) from the same Blue Marble mosaic used
220 for determination of the LSM. The upstream catchment size is suggested to provide an estimate
221 of the transport and deposition of sediments and highlights areas of sediment accumulation,
222 whereas the reflectance serves as a proxy for highly erodible surfaces such as playas and dunes.
223 The value for the SSM is based on a scale from 0 – 1, with the Bodélé Depression in Chad
224 regarded as the most emissive source with a maximum value of 1 (Parajuli & Zender, 2017).

225 3.2 PI-SWERL dust-emission measurements

226 Dust-emission measurements from the PI-SWERL instrument were used to measure the
227 potential for dust flux from different desert surfaces, with the PI-SWERL now being a widely
228 used technique (e.g., Bacon et al., 2011; Etyemezian et al., 2007; Goossens & Buck, 2009;
229 Sweeney et al., 2008, 2011). The specific methodology and test parameters for the PI-SWERL
230 are presented in von Holdt et al. (2017). The dust emission flux (E_f , $\text{mg m}^{-2} \text{s}^{-1}$) was calculated
231 using the following equation from Sweeney et al. (2011):

$$232 \quad E_f = \frac{\sum_{begin,i}^{end,i} C \times F}{(t_{end,i} - t_{begin,i}) \cdot A_{eff}}, \quad (1)$$

233 where C is the dust concentration (mg m^{-3}) of PM_{10} (Particulate matter $<10 \mu\text{m}$), F is the flow
234 rate of air through the chamber (L s^{-1}), A_{eff} is the test area underneath the PI-SWERL annular
235 ring (m^2), and t is the time (s) at the beginning ($t_{begin, i}$) and end ($t_{end, i}$) of the RPM step test
236 level, i (Sweeney et al., 2011). In order to take measurements at sites of known dust emission,
237 the PI-SWERL was deployed at a total of 17 sites identified from Landsat (von Holdt et al.,
238 2017) (Figure 1 a). A further three sites were tested representing low emission surfaces, i.e. the
239 gravel pavements within stony systems and sand dunes within aeolian systems. These non-
240 emissive sites were included in the study to obtain the full range of emission potential and for
241 purposes of the regression analysis. The stone pavements chosen as test sites were selected
242 based on the presence of a vesicular A soil horizon (Av) (McFadden, 2013, Sweeney et al.,
243 2013) as these horizons are indicators of dust activity, predominantly as inputs to soils.

244 At each site a 10 m tape marked at 1 m intervals was laid out as close as possible to the co-
245 ordinates of emission source points determined from Landsat. For consistency, transects were

246 all laid out perpendicular to the direction of the north-east regional wind (the Bergwind) that is
247 responsible for much of the large-scale dust emission in the Namib (Vickery & Eckhardt,
248 2013). All 17 sites underwent visual confirmation on the ground of the surfaces identified as
249 potentially emissive. The site was assessed, and a final *in situ* decision was made regarding the
250 placement of the transect so as to include all the different surface types that were apparent
251 locally. At each location of testing, 3 to 10 individual runs were made at metre intervals along
252 a 10 m linear transect with the number of test runs dependent on the homogeneity of the
253 surfaces within the transect and variability of the emission flux results.

254 A further decision regarding the number of runs to perform was made based on the PI-SWERL
255 results at the time of testing. Floodplain terraces present within alluvial systems composed of
256 silt crusts with variable amounts of sand for saltation and nebkhas situated on the terraces as
257 well as loose erodible material present in between the silt crusts proved to be highly variable
258 and as a result, 10 transects were done on these terraces. In contrast, sand dunes within the
259 aeolian systems were relatively uniform in emission potential and so fewer measurements were
260 carried out on these surfaces and were largely for exploratory purposes (Table 1). Dune sand
261 deposits have not been identified as significant point source emitters from Landsat but have
262 been identified as low intensity dust sources covering large areas and hence a potentially
263 appreciable source of dust (Bullard et al., 2004; Crouvi et al., 2008; Strong et al., 2010).

264 The PI-SWERL measurements were classified according to the individual erosional surfaces
265 that were being tested. The individual surfaces were then aggregated first to a landform scale
266 and lastly to a landscape scale (Table 1). Details of the PI-SWERL test sites and landform
267 classifications are given in Table S1 of the Supporting Information. We used a mixed effects
268 model to investigate the relationship of dust emission potential within each spatial scale (suited
269 to unbalanced replicates) with catchment identity set as a random effect (as in Sankey et al.,
270 2011) followed by an analysis of variance for the fixed-effects terms (Table 1 categories for
271 Landscape, Landform and Surface respectively). Data at all scales were log-transformed before
272 model runs to satisfy the assumption of the normality of the residuals. All models and
273 significance testing were performed in R 3.4.1 (R Development Core Team, 2017) using the
274 ‘nlme’ package (Pinheiro et al., 2018). A threshold *p*-value of <0.05 was regarded as
275 significant. A determination of spatial autocorrelation using the Moran’s I test statistic was
276 performed with the ‘ape’ package (Paradis & Schliep, 2018) in R 3.4.1. at all spatial scales.

277 Table 1 Breakdown of the categories of geomorphology considered in the statistical analysis^a

SURFACE ($<10^{-1}$ m) (n =individual PI-SWERL measurements) $n=128$	Loose Erodible Material (LEM) (25)	Crust: high saltators (22)	Crust: medium saltators (12)	Crust: no saltators (31)	Low % gravel (17)	High % gravel (9)	Salt crust: with and without saltators (12)
LANDFORM (10^1 - 10^2 m) (n =PI-SWERL transects) $n=17$	Active channel (1) Terraces (10)	Drainage channel (2) Pavement (4)			*	*	
LANDSCAPE (10^3 m) (n =landforms sampled) $n=12$	Alluvial systems (5)	Stony systems (4)			Aeolian systems (1)	Ephemeral lake systems (2)	
CATCHMENT	Kuisseb, Omararu, Huab, Ugab						

279 ^a Increasing spatial scale of enquiry moving down the table. Catchment was set as random
 280 effect across the Namib Desert study region. ‘Aeolian systems’ and ‘Ephemeral lake systems’
 281 landscapes were not analysed at Landform scale due to insufficient number of sample points
 282 and sampling conditions.

283

284 3.3 Characterization of surface properties and Boosted Regression Tree (BRT) analysis

285 A BRT model was used to identify the most relevant variables that controlled surface
 286 erodibility using the surface properties at each PI-SWERL field testing site. This analysis was
 287 performed following Elith et al. (2008) using the ‘*dismo*’ package (Hijmans et al., 2016) in R
 288 with a learning rate of 0.005 and a tree complexity of 5. E_f (Equation 1), representing the overall
 289 surface erodibility, was used as the response variable and specific predictor variables included
 290 compressive and shear strength to quantify the degree of crusting, soil moisture content,
 291 particle size and elemental composition to assess the influence of mineralogy. For those
 292 properties tested in the laboratory, one surface sample was taken with a flat spade to a depth of

293 0.02 m directly next to each PI-SWERL run at the time of testing. Further details of the BRT
294 analysis are given in Section S1 of the Supporting Information.

295 The degree of consolidation or crusting of the surface was assessed in the field by measuring
296 the compressive and shear strength of the surface. Unconfined compressive strength was
297 measured on the test surface at each PI-SWERL run site using a Pocket Soil Penetrometer H-
298 4195 and shear strength using a Torvane H-4212 pocket shear vane (Humboldt Mfg. Co.,
299 Illinois, USA). At each site a minimum of three measurements of both compressive and shear
300 strength of the surface were taken. If a large difference in individual measurements was
301 encountered, additional measurements were taken to increase representativeness of
302 measurement.

303 The near-surface volumetric soil moisture content at the time of PI-SWERL sampling was
304 measured in the field with a Delta-T Devices ML3 ThetaProbe soil moisture sensor. A
305 minimum of three measurements were taken at each PI-SWERL measurement site by inserting
306 the probe to just below the soil surface. We note that the probe was designed to be fully inserted
307 into the soil medium (to a depth of 0.06 m), producing values for a deeper soil volume. Our
308 non-standard application of the instrument (not inserting it fully into the soil) is intended to
309 provide a relative measure of near-surface soil moisture, but has not been vetted through further
310 gravimetric measurements. The ThetaProbe was therefore inserted only into the top 0.02 m.
311 For this study, establishing a gauge of soil moisture as close as possible to the surface (which
312 has a strong influence on erosion potential (Wiggs et al., 2004) outweighed specific
313 quantification of soil moisture level.

314 For particle size analysis, all the samples were air dried at 25°C to a constant weight and sieved
315 to 1 mm. The >1 mm split was further sieved to determine the coarse sand and gravel fractions.
316 The <1 mm split was used to determine the particle size distribution by laser diffraction using
317 a Malvern Mastersizer 2000 attached to a Hydro 2000G dispersion unit. The samples were cone
318 and quartered to obtain a representative sample and placed in a tap water solution overnight,
319 shaken for half an hour and again for half an hour the next day before introduction to the
320 dispersion unit and further sonicated for 180 seconds prior to measurement. Particle size
321 statistics including texture classes, modes, kurtosis, skewness and sorting were calculated using
322 Gradistat software (Blott & Pye, 2001).

323 The gravel concentration in the surface sediments determined by sieve analysis served as a
324 proxy for gravel-cover density. These results were confirmed by an unsupervised classification

325 of a photograph of each surface with a gravel cover performed in Erdas Imagine 2015-2016
326 (Leica Geosystems, Atlanta, Georgia, USA). Gravel cover densities of <30% were classified
327 as low gravel cover and densities of >30% as high gravel cover. This distinction was chosen
328 based on the analysis by Wang et al. (2012), who found that dust emission increases with
329 increasing gravel cover up to a density of 30%, after which dust emission decreases.

330 Finally, milled samples of the <1 mm soil fraction were analyzed in the laboratory for Mg, Al,
331 S, Cl, K, Ca, Ti, Mn and Fe. with a Spectroscout energy-dispersive X-ray Fluorescence (XRF)
332 analyzer (SPECTRO Analytical Instruments, Kleve, Germany). The instrument was calibrated
333 with a certified standard GBW07312 (National Research Centre for CRMs, Beijing, China) for
334 which technical concentrations were obtained from NOAA Technical memorandum NOS
335 ARCA 68 (1992).

336 **4 Results**

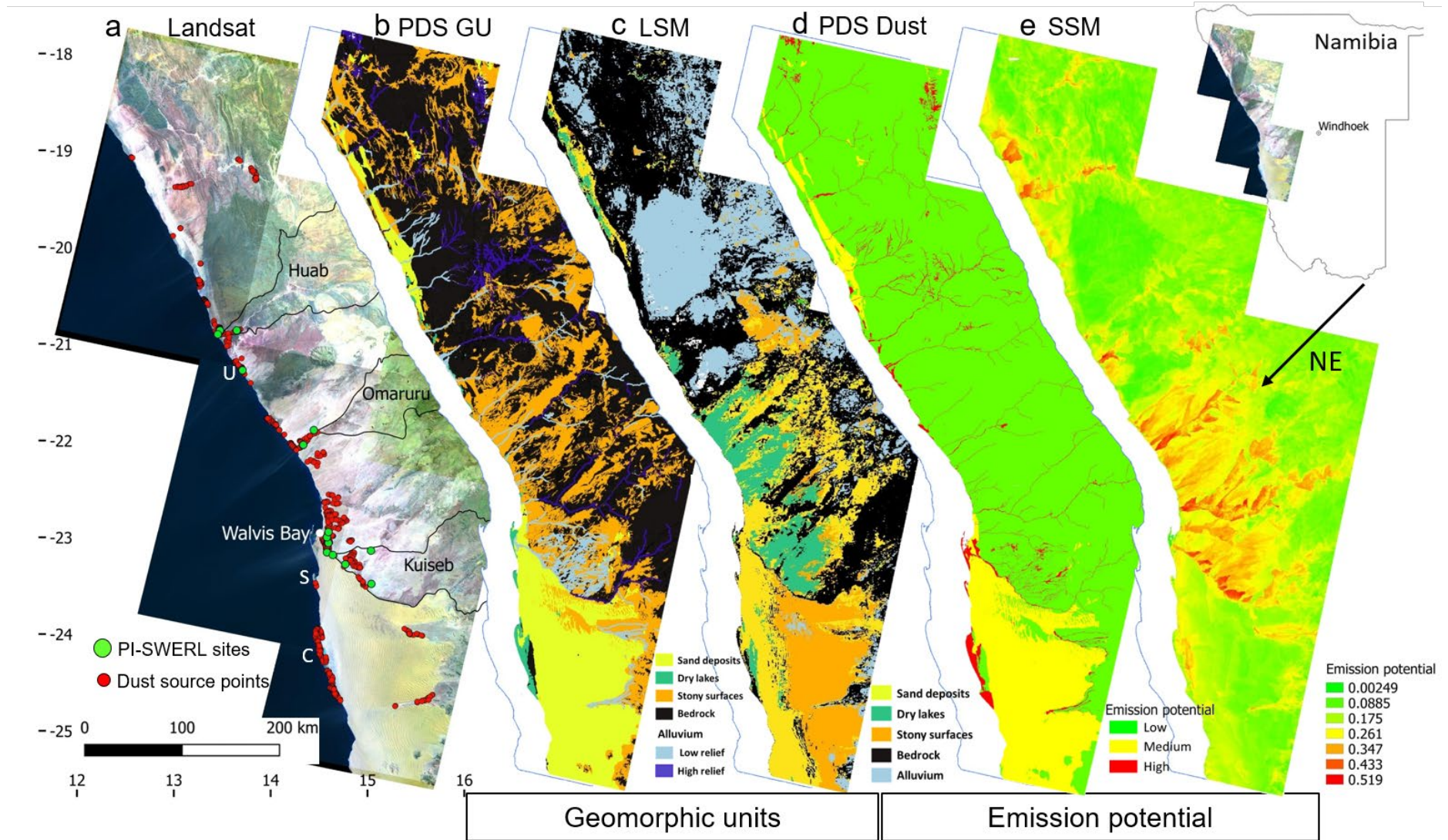
337 4.1 Dust emission scheme mapping and Landsat-derived dust source points

338 Stony systems and bedrock cover extensive areas of the Namib study area when mapped
339 according to the PDS classification (Figure 1 b). Two extensive aeolian systems of the Namib
340 Sand Sea in the south and the Skeleton Coast dunefield in the north account for the second
341 largest portion of the land area. In the present study, the PDS classification was applied to the
342 2289 dust source points observed from Landsat imagery (1990-2016) by von Holdt et al. (2017)
343 (Figure 1; Figure 2). Overall, the ephemeral lake systems and alluvial systems cover a very
344 small proportion of the study area (2% of area) but contribute just over three-quarters of
345 observed source points (77% of plumes) (Figure 2). In contrast, are stony systems (27% of
346 area, 22% of source points) and aeolian systems (15% of area, 0.5% of source points), which
347 cover large areas, but contain fewer point sources of dust emission. Additional details of the
348 landform classification for the dust source points performed in the present study are given in
349 section S2 of the Supporting Information.

350 The representation of the landscape according to the LSM (Figure 1 c) is noticeably different
351 from the PDS, with large areas of bedrock and stony systems classified as either alluvial system
352 or ephemeral lake. In addition, a large part of the Namib Sand Sea is classified as stony system,
353 an issue noted by Parajuli & Zender (2017). Relatively small areas of a landscape are
354 responsible for most of the dust emission (e.g. Bullard et al., 2008; Gillette, 1999; Lee et al.,
355 2009) which is evident when assigning a level of dust emission potential to the PDS land

356 surface classification following Bullard et al. (2011) (Figure 1 d). Alluvial systems and
357 ephemeral lake systems are the highest potential emitters, aeolian systems have low to medium
358 dust emission potential and stony systems and bedrock are low potential emitters. The colours
359 assigned to low, medium and high emission potential categories follow the colour scheme used
360 in the SSM (Figure 1 e) by Parajuli & Zender (2017). The SSM highlights the elevated potential
361 of alluvial systems to emit dust, but when mapped, results in a more extensive alluvial coverage
362 than represented by the PDS scheme.

363

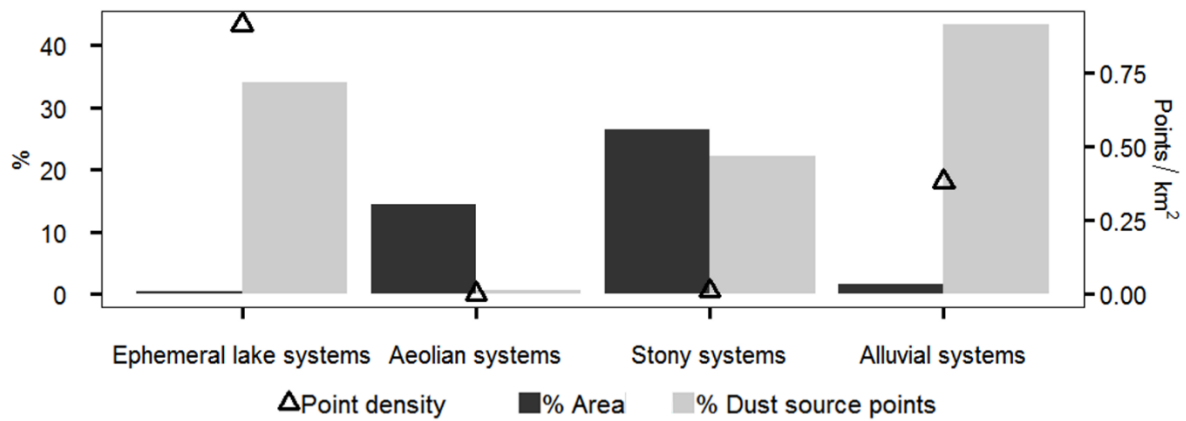


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366 **Figure 1 Geomorphology and dust emission potential mapping of the Namib Desert area covered in this study. (a) Landsat false-colour image**
367 **showing the seven tiles included in the Landsat source point analysis, and the 2289 Landsat points from von Holdt et al. (2017), key river catchments**
368 **and PI-SWERL testing sites, (b) PDS GU geomorphic land surface classes per Bullard et al. (2011), (c) Land Surface Map (LSM) of Parajuli &**
369 **Zender (2017), (d) dust emission potential according to PDS emission categories (PDS Dust), (e) Sediment Supply Map (SSM) showing dust emission**
370 **potential based on surface reflectance on a unitless scale from 0 to 1 with the maximum value equated to the Bodélé Depression (Parajuli & Zender**
371 **(2017). In (a), U: Ugab pan complex, S: Sandwich Harbour, C: Conception Bay. Direction of predominant north easterly dust-producing winter**
372 **wind (Bergwind) is indicated by black arrow in (e).**

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376 **Figure 2 Areal extent of geomorphic landscape classes (% of total area) and frequency of dust**
377 **source points within them (expressed as % of total number, as well as points per km²) identified**
378 **through Landsat analysis of the Namib Desert. Ephemeral lake systems (which include playa**
379 **and sabkha pans) have the lowest extent within the study area (850 km²), but show the highest**
380 **density of source points. The stony systems have the highest areal coverage (45,000 km²), but**
381 **show a low density of source points. Alluvial systems have the highest number of source points**
382 **overall (43%) but cover 4% of the study area. Aeolian systems cover 15% of the study area but**
383 **were responsible for <1% of the source plumes identified.**

384

385 4.2 Measured emission fluxes

386 The surface scale analysis of the PI-SWERL measurements illustrates the inherent variability
387 of dust emission at a sub-landform scale (Figure 3). At this scale, the most emissive undisturbed
388 surfaces occurred where loose erodible material (LEM) was present. The presence of such
389 material was particularly associated with the presence of small nebkha dunes interspersed
390 between crusted fluvial deposits within the valley fill terraces and within the drainage channels
391 of the stony systems. In Figure 3, a distinction between the crusted surfaces present in the
392 channels or on the terraces could be made based on the relative presence of saltators determined
393 by inspection of the surface before a PI-SWERL test. Figure 3 indicates that LEM-dominated
394 surfaces (geometric mean: $0.3188 \text{ mg m}^{-2} \text{ s}^{-1}$) and those crusts with abundant sand for saltation
395 (geometric mean: $0.342 \text{ mg m}^{-2} \text{ s}^{-1}$) were significantly more emissive than the other surface
396 types (p value < 0.001) (summary statistics in Table 2). Pavement surfaces with varying
397 densities of gravel were found predominantly within the stony systems and in some river
398 terraces. The low-density stone pavement (gravel cover $< 30\%$) was significantly more emissive
399 (geometric mean: $0.02004 \text{ mg m}^{-2} \text{ s}^{-1}$) than the surfaces with a high density of gravel cover
400 ($> 30\%$). High density gravel surfaces (geometric mean: $0.0022 \text{ mg m}^{-2} \text{ s}^{-1}$), crusts with no
401 saltators (geometric mean: $0.0046 \text{ mg m}^{-2} \text{ s}^{-1}$) and salt crusts (geometric mean: 0.0008 mg m^{-2}
402 s^{-1}) were the lowest emitters. All p values for significance tests are reported in the Supporting
403 Information Tables S3 to S5.

404 Aggregating the observed emission fluxes within the landscape scale classes found in dust
405 emission potential schemes illustrates the problematic nature of representing sub-landform
406 scale variability at a larger scale (Figure 4). The greatest amount of variability was present in
407 the stony and alluvial systems and when aggregated to the landscape scale, the geometric means
408 for these two classes were not significantly different (Figure 4 a). Notably, the lake systems
409 are significantly different (p value = 0.040) and consistently showed low emissivity during the
410 time they were tested (geometric mean: $0.0022 \text{ mg m}^{-2} \text{ s}^{-1}$). The ephemeral lake systems tested
411 included the Huab playa and Ugab sabkha (Figure 1 a), where significantly less dust was
412 emitted than the other three geomorphic landscape units as quantified by the PI-SWERL
413 (alluvial systems geometric mean: $0.0379 \text{ mg m}^{-2} \text{ s}^{-1}$ and stony systems geometric mean:
414 $0.0102 \text{ mg m}^{-2} \text{ s}^{-1}$). Aeolian systems were not included in the landscape-scale analysis due to
415 insufficient sample size, but had a geometric mean of $0.0640 \text{ mg m}^{-2} \text{ s}^{-1}$ based on the individual
416 PI-SWERL measurements (Table 2).

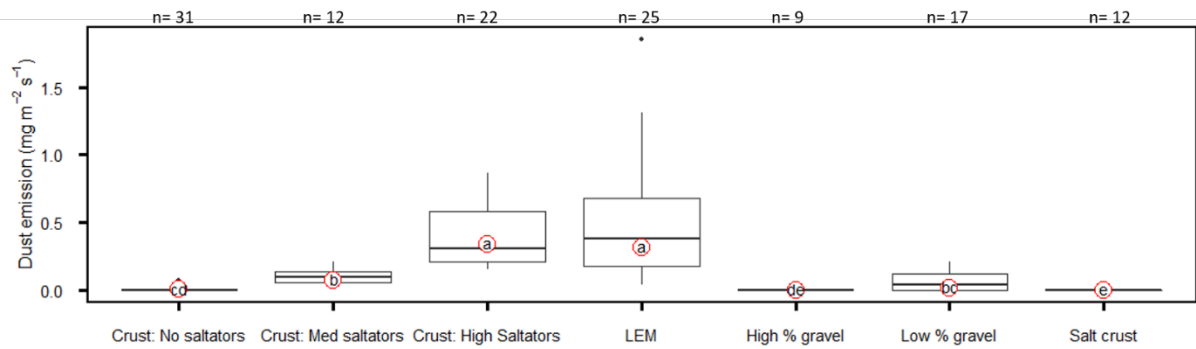
417 The variability in the alluvial and stony systems was further resolved by looking at distinct
418 landforms present within these two broad landscape classes (Figure 4 b and c). In the case of
419 the stony systems class, a fundamental distinction could be made between stone pavement
420 surfaces dominated by the presence of coarse lag gravel, and portions of pavement where micro
421 drainage channels (c. 0.1 m deep) were found (Figure 4 d and e). In turn, the low relief alluvial
422 class could also be divided between portions of ephemerally active river channel and valley fill
423 terraces, the latter situated above the channel (Thomas et al., 2017) (Figure 4 f and g). The river
424 valley fill terraces (within alluvial systems) were on average the most emissive landform
425 (geometric mean E_f : 0.0651 mg m⁻² s⁻¹), followed by the stony systems exhibiting drainage
426 channels (geometric mean: 0.0318 mg m⁻² s⁻¹). The gravel pavements (geometric mean: 0.0075
427 mg m⁻² s⁻¹) and active river channels (geometric mean: 0.0082 mg m⁻² s⁻¹) were less emissive.
428 In terms of statistical separation, however, only the stony pavements had a lower emission rate
429 than the alluvial terraces (p -value = 0.00769).

430 The rapid, multi-replicate PI-SWERL testing allows the spatial variability in emission flux
431 from a given surface, landform or landscape to be measured (King et al., 2011; Sweeney et al.,
432 2011). The same crust within a 10 m transect can be largely non-emissive (0.003 mg m⁻²s⁻¹) in
433 the absence of available sand for saltation, but highly emissive (0.646 mg m⁻²s⁻¹) where an
434 abundant supply of saltators is present. Emission rates generated by the PI-SWERL testing of
435 surfaces reflect the relative presence of only those saltators under the instrument footprint,
436 resulting in non-emissive runs on crust where no saltators are present. However, river terraces
437 surrounded by an abundant supply of sand will undergo widespread bombardment by saltation
438 during a high friction velocity wind event. In such circumstances, it is possible that the entire
439 transect will become highly emissive under the continued bombardment of the available
440 saltators and stockpiles of LEM dispersed between the terraces. To test the degree to which
441 individual measurements were spatially autocorrelated, a Moran's I test was performed for
442 individual measurements within a transect and were not found to be spatially autocorrelated
443 (for example $p = 0.0705$ using Moran's I for Huab transect 2 and $p = 0.090$ for Kuiseb 5)
444 indicating that the sampling density was adequate, and autocorrelation is not relevant at the
445 surface scale analysis. A higher sampling density is not possible given the size of the PI-
446 SWERL.

447 4.3 Emission fluxes and relation to land surface classification schemes

448 The PI-SWERL provides a relative quantification of dust emission rates from the surface,
449 against which the emission potential of different geomorphic classes in surface classification
450 schemes (PDS, SSM/LSM) can be compared. Comparing the geometric mean of measured dust
451 emission of the PI-SWERL transects across landforms against the SSM index values for the
452 location of each PI-SWERL transect provides a means to assess and contextualise the SSM
453 values (Figure 5). Also represented in Figure 5 is the land surface classification as per the PDS
454 scheme by Bullard et al. (2011) and the LSM classification by Parajuli et al. (2014).
455 Determination of the classification between the two different schemes differs considerably, for
456 instance, with LSM classifying two out of the 20 transect locations as bedrock, while PDS
457 identified them as either dry lake or alluvial systems (Figure 5). Elsewhere, LSM was found to
458 classify PDS alluvial systems as bedrock with sediment and stony systems as Playa/Sabkha.
459 The PI-SWERL results do not show a clear relationship between the measured dust emission
460 and the SSM values. The SSM values for the entire study area range from a minimum of 0.002
461 to a maximum of 0.519 with a mean of 0.187, with the peak geometric mean transect emission
462 rate ($0.2191 \text{ mg m}^{-2} \text{ s}^{-1}$) corresponding to a moderate SSM value (0.25) (Figure 5).
463 Furthermore, a wide range of emissivity (0.002 to $0.2191 \text{ mg m}^{-2} \text{ s}^{-1}$) is seen in the narrow
464 range of SSM values between 0.23 to 0.27. This range covers the LSM categories of stabilised
465 sand deposit, sand deposit on bedrock and bedrock, but is more appropriately classified as
466 predominantly alluvial system and some stony system according to the PDS map. The highest
467 SSM value for the PI-SWERL test sites was 0.46, which corresponded to a stony system with
468 an emission value of $0.0127 \text{ mg m}^{-2} \text{ s}^{-1}$. Other high SSM values (>0.3) mostly occurred within
469 alluvial systems, with measured emission values varying widely between 0.008 and 0.12 mg
470 $\text{m}^{-2} \text{ s}^{-1}$, and the lowest emission flux value in this range associated with the active channels.
471 The locations of the dust source points identified by von Holdt et al. (2017) with Landsat
472 imagery in Figure 1 (a) have SSM values with a range of 0.078 to 0.508 and a mean of 0.245.

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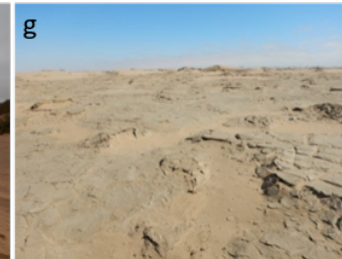
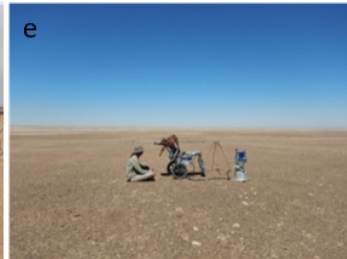
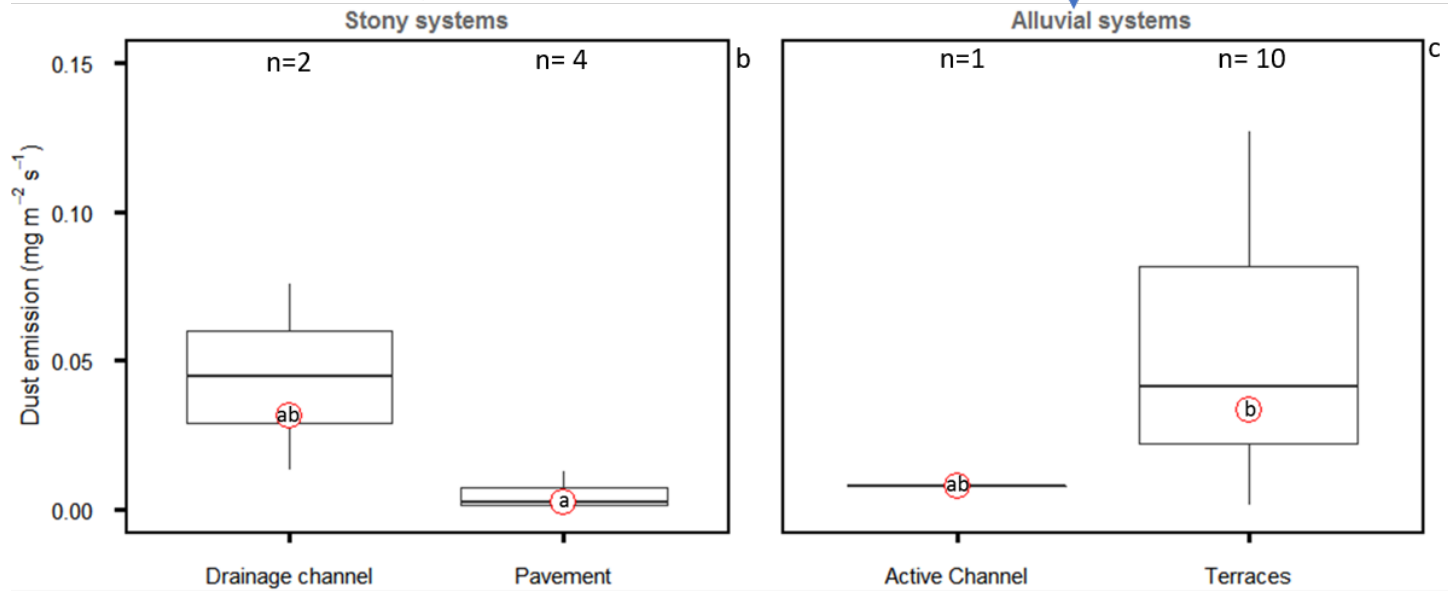
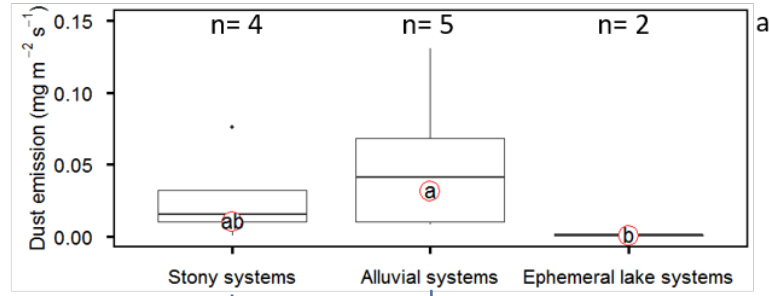
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Figure 3: Dust emissions from surface categories within all landscapes. The loose erodible material (LEM) surface type consists of unconsolidated sediments and is found in ephemeral drainage channels of stony systems and on terraces of alluvial systems between silt crusts and nebkhas (Figure 4, left and right photos, respectively). The measurements carried out in the aeolian system were also included in the LEM category. The crust surface class occurs in alluvial systems and was subdivided based on abundance of sand available for saltation (none, medium, high). Gravel surfaces found both in stony systems in pavements and alluvial systems in terraces were subdivided based on the density of gravel cover (Low % with <30% gravel cover and High % with > 30% gravel cover). Salt crusts were found within ephemeral lake systems with and without saltators present. Letters indicate significant difference and are plotted at the geometric mean of each surface category.



488 **Figure 4 Dust emission determined from PI-SWERL measurements at landscape scale for stony, alluvial and ephemeral lake systems (a). Stony**
489 **systems consisted of two landform categories: stone pavements and pavement intersected by ephemeral drainage channels (b). Similarly, alluvial**
490 **systems contained two landforms: active river channels and floodplain terraces (c). Corresponding photographs of each landform are shown at the**
491 **bottom (d to g). Circled letters indicate statistical difference determined and are plotted at the geometric mean value for each distribution. River**
492 **terraces proved to have the highest dust emissions in alluvial systems, while no significant difference was observed between drainage channels and**
493 **pavement within stony landscapes.**

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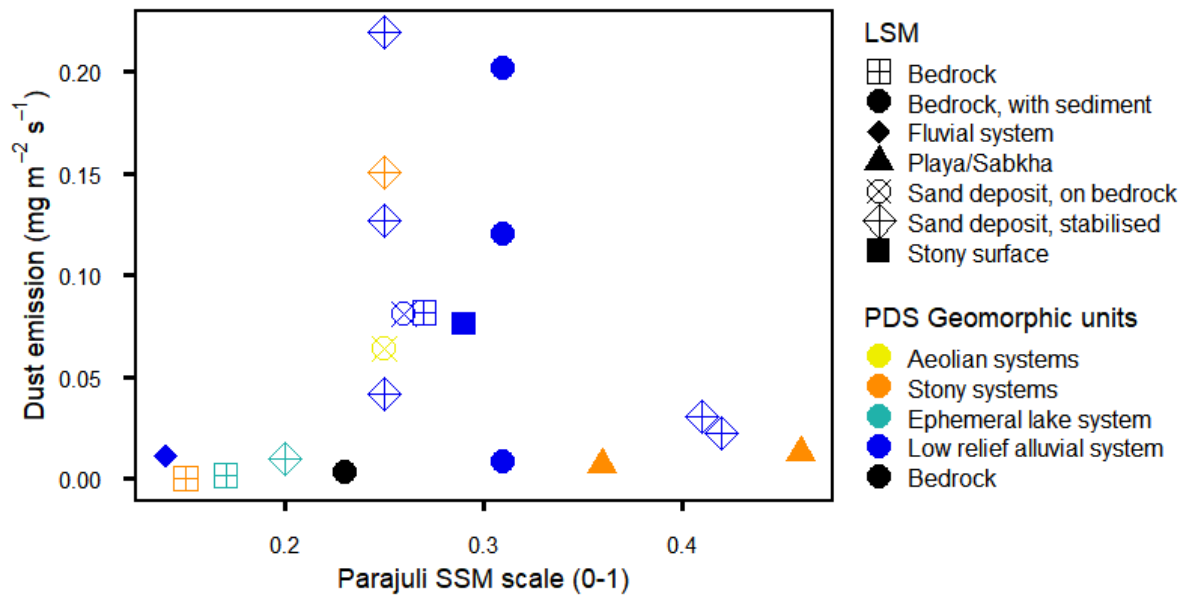
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498 Table 2 Summary statistics for PI-SWERL dust flux measurements at the Landscape,
 499 Landform and Surface scale assessments
 500

	n ^a	n ^b	Geometric mean (mg m ⁻² s ⁻¹)	CI low ^c (mg m ⁻² s ⁻¹)	CI high ^d (mg m ⁻² s ⁻¹)	Geo SD (mg m ⁻² s ⁻¹)	Min ^e (mg m ⁻² s ⁻¹)	Max ^f (mg m ⁻² s ⁻¹)
Surfaces (n = 128 individual PI-SWERL measurements)								
Low % gravel	17	17	0.0204	0.0076	0.0566	8.544	0.0005	0.2129
High % gravel	9	9	0.0022	0.0009	0.0052	4.378	0.0004	0.0155
Loose erodible material	25	25	0.3188	0.2143	0.4861	2.898	0.0417	1.854
Crust: no saltators	31	31	0.0046	0.0027	0.0071	4.016	0.0003	0.0704
Crust: med saltators	12	12	0.0855	0.0578	0.1359	2.258	0.0102	0.2097
Crust: high saltators	22	22	0.3418	0.2690	0.4356	1.730	0.1575	0.8649
Salt crust: with and without saltators	12	12	0.0008	0.0002	0.0026	8.146	0.00006	0.0210
Total n	128	128						
Landform (n = 17 PI-SWERL transects)								
Terraces	10	75	0.0651	0.0376	0.1194	2.680	0.0111	0.2191
River channel*	1	9	0.0082	0.0014	0.0491	10.258	0.0009	0.1689
Pavement	4	11	0.0075	0.0010	0.0436	10.416	0.0006	0.1501
Drainage channel	2	13	0.0318	0.01332	0.0759	3.422	0.0133	0.0758
Total n	17	108 ^g						
Landscape (n = 12 landforms)								
Alluvial systems	5	84	0.0379	0.0179	0.0881	2.848	0.0082	0.1310
Stony systems	4	24	0.0102	0.0019	0.0726	7.468	0.0006	0.0759
Aeolian systems*	1	3	0.0640	0.0406	0.1001	1.199	0.0534	0.0767
Ephemeral lake systems	2	17	0.0022	0.0005	0.0094	7.479	0.0005	0.0094
Total n	12	128						

501 ^a Sample size n after aggregation of individual PI-SWERL measurements to relevant scale of enquiry
 502 ^b Sample size n using individual PI-SWERL measurements
 503 ^c 95% confidence interval below the mean ^d 95% confidence interval above the mean
 504 ^e Minimum emissions from unit/surface
 505 ^f Maximum emissions from unit/surface
 506 ^g Landform assessment does not include Aeolian systems (n=3) and Lake systems (n=17)
 507 *Summary statistics calculated with individual measurements as insufficient n at aggregated level (n=1)
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514 **Figure 5** Dust emission flux measured with the PI-SWERL compared to the SSM (Sediment
 515 **Supply Map)** value for each of the transect sample sites. The legend also indicates the land
 516 **surface classification according to the LSM (Land Surface Map) (by symbol shape), and**
 517 **secondly, as mapped using the land surface classes proposed by the PDS (Preferential Dust**
 518 **Scheme) by symbol color.**

519

520 4.4 Predictors of emission rate as determined by Boosted Regression Tree analysis

521 The BRT model produced the following variables as the most important predictors for dust
522 emitted during the PI-SWERL runs: gravel cover (%), moisture content (%), kurtosis, very
523 coarse silt fraction (%), very fine sand (%), fine sand (%), compressive strength (kg m^{-2}), Ca
524 (%), Mg (%) and S (%). The relative contribution of each variable to the model is given as a %
525 and the partial dependence plots (Figure 6) provide the relationship between the variables and
526 the measured dust flux when all other variables are held constant. The trend in the plots is
527 informative, rather than actual values, with increasing partial dependence values indicating
528 increased dust emission and vice versa. A sudden change indicates a critical threshold at which
529 the dust emission flux changes. Taken together, the significant predictor variables identified
530 with the BRT explain 70.8% of the deviance in the dust flux measured with the PI-SWERL.

531 Based on the BRT analysis, soils layers with a content of very coarse silt above 5% and a very
532 fine to fine sand content between 10 and 20%, resulting in a platykurtic particle size
533 distribution, should indicate areas with potentially increased emission potential. In addition,
534 the density of gravel cover results in an increase in roughness and bed armoring which appears
535 to exert a significant influence in reducing emission potential when gravel content is 15% or
536 above. Moisture has long been regarded as a primary control on dust emission (e.g. Ishizuka et
537 al., 2005; McKenna-Neuman & Nickling, 1989; Munkhtsetseg et al., 2016) and emerges as a
538 primary predictor. Calcium and magnesium were also identified as important elements
539 potentially due to the effect that carbonate minerals have on the erodibility of a crusted soil,
540 with some suggesting these minerals will act to strengthen crusts by acting as a binding agent
541 (Gillette et al., 1982) and others contending that calcite offers very little resistance to abrasion
542 (Pye and Tsoar, 1990). Our data seem to support a reduction in dust flux with increasing Ca
543 and Mg content.

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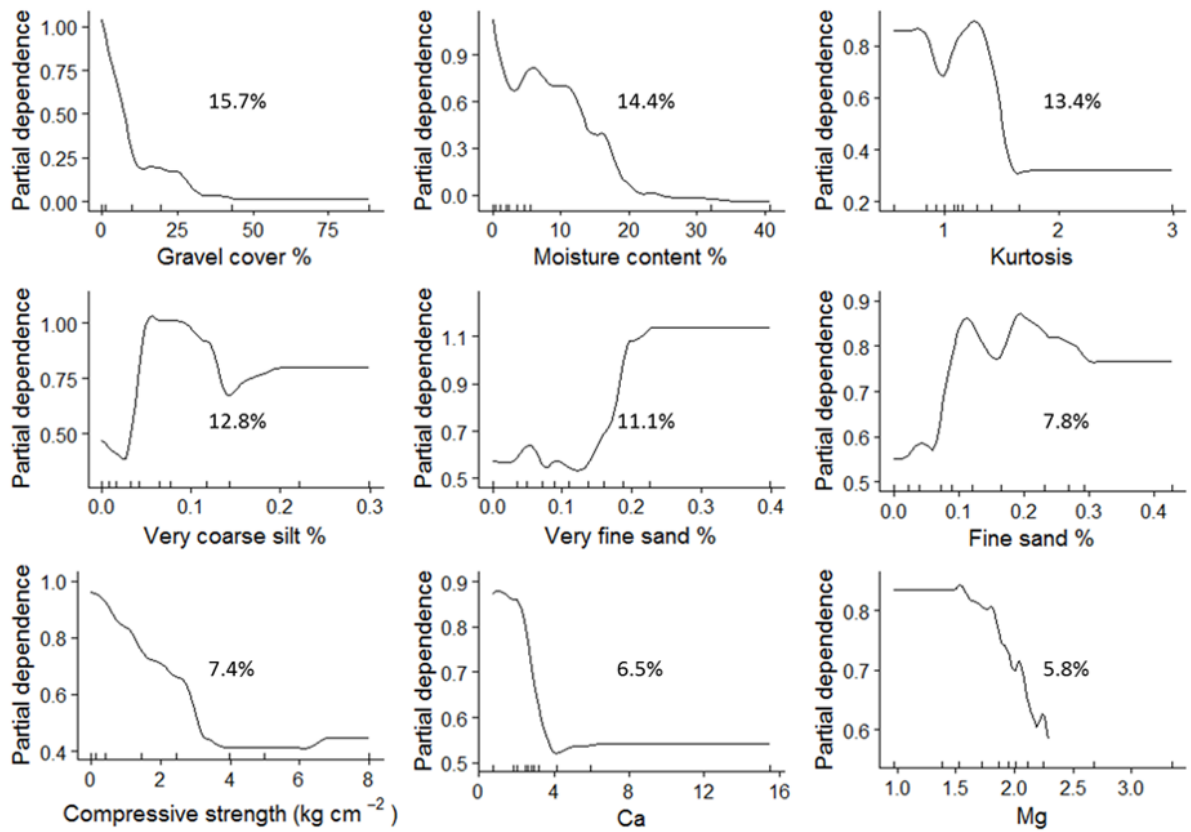
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551 **Figure 6 Partial dependence plots depicting the relationship between dust emission and each**
 552 **significant surface variable. The trend rather than the actual values is the important feature in**
 553 **each plot. Increasing partial dependence values indicates an increase in dust flux and decreasing**
 554 **values the opposite. Percentage values reported at the centre of each panel are the relative**
 555 **influence of variables for predicting dust emission.**

556

557 5 Discussion

558 The acquisition of landform scale dust source data achieved here allows for the evaluation of
559 the PDS and LSM classification schemes. It also allows an assessment of the newly generated
560 landscape-scale SSM product to characterise dust emission potential within the Namib Desert.
561 The Namib represents an ideal region for such an investigation as it is host to a variety of
562 actively emitting surfaces (Vickery et al., 2013; von Holdt et al., 2017). Classifying the dust
563 source points identified with Landsat by von Holdt et al. (2017) according to the PDS at a
564 landscape scale indicates that dust emission from the Namib Desert is spatially highly
565 concentrated, with relatively high densities of plumes found to originate from the alluvial
566 systems (0.4 points km²) and dry lakes (0.9 points km²) compared to aeolian (0.001 points
567 km²) and stony systems (0.011 points km²) (Figure 2). Mapping the land surface classes of the
568 Namib Desert according to the PDS at landscape scale shows the limited extent of ephemeral
569 lakes (playas and sabkhas) as dust producing areas (2% of the total study area) which reflects
570 the hotspot nature of dust production in these landscapes noted by Gillette (1999).

571 The advantage of the PDS map is that it can represent the landscape in detail because of the
572 high-resolution, quality controlled geomorphic attribution of the surface that comes from user-
573 defined mapping. The disadvantage of this scheme is that it requires certain inputs to map the
574 landscape, which are not consistently available for all areas, and its critical requirement for
575 land surface classes or geomorphic units to be identified and created, which is prone to
576 subjectivity. The PDS mapping has only been performed for limited areas (e.g. Baddock et al.,
577 2011; Bullard et al. (2011); Lee et al., 2012) and creating a global PDS map remains a
578 challenge. Using the LSM developed by Parajuli et al (2014) to map the study area results in
579 an overestimation of the dust emitting alluvial and dry lake areas (Figure 1 c). This is due to
580 the misclassification of several land surface classes which occurs as a result of a supervised
581 technique with training classes based on the spectral signature of the MENA (Middle East and
582 Africa) region (Parajuli & Zender, 2017). The difficulties of such an automated approach
583 demonstrates the problems of attempting to create a global geomorphology classification map.
584 Even though a global land surface classification or geomorphology map would provide a
585 valuable input to the representation of dust emission, the use of region specific training classes
586 should be exercised with caution, especially when based on spectral data. A further limitation
587 of a qualitative geomorphic mapping scheme, such as the PDS and LSM, involves the
588 representation of a quantified dust emission potential.

589 A quantified representation of the dust emission potential for different land surface classes at
590 a landscape-scale in raster format is necessary to incorporate these schemes into dust cycle
591 models. For the PDS this has not been achieved and each class is assigned a qualitative
592 categorical indicator based on inferred emission potential, although the scheme has been tested
593 against long-term frequency of dust observation in the Chihuahuan Desert (Baddock et al.,
594 2016). With PDS, emission information is required to discriminate between relative emission
595 potential from different regions that act as dust sources. A quantification of the dust emission
596 potential of the LSM land cover categories was attempted by Parajuli et al. (2014) using a
597 correlation between ERA-Interim wind speed at 10 m height and MODIS deep blue AOD at
598 550 nm. The authors point out that a disadvantage of this approach is the difference in scale
599 between the high-resolution land cover map and the coarser ($1^\circ \times 1^\circ$) correlation map which
600 results in a disconnect between land cover and the emission potential assigned to them. The
601 location of the major Namib Desert dust sources in the low-relief terminal stages of the rivers
602 and coastal lake systems (pans and sabkhas) poses difficulties for identification of these sources
603 based on techniques relying on aerosol loadings. Furthermore, the use of atmospheric aerosol
604 loading estimates, such as MODIS AOD or TOMS AI to locate dust sources in the Namib
605 Desert may well have specific limitations. For example, detection of dust over bright desert
606 surfaces using ultraviolet, visible or thermal infrared wavelengths can be problematic (Baddock
607 et al., 2009; Hsu et al., 2004; Resane et al., 2004). In contrast, MODIS Deep Blue (MODIS
608 DB) can only be retrieved over bright surfaces and is of limited use over dark ocean surfaces,
609 while TOMS AI is known to not detect dust from the Namib Desert at low altitude near the
610 coast (Mahowald & Dufresne, 2004). A consequence of this suite of limitations is that creation
611 of an erodibility map derived from supervised classes established in a different region together
612 with reliance on emission quantification from satellite retrieved aerosol loadings would likely
613 result in a number of dust producing areas, such as the Namib Desert, being underestimated.
614 Field based studies that include PI-SWERL emission measurements from intensely-sampled
615 regions can provide relative dust fluxes, as well as indications of variability, that can serve as
616 inputs for the quantification of dust potential schemes such as the PDS and LSM (Table 2).

617 The recently proposed SSM provides a global landscape-scale erodibility map with a
618 quantification of dust emission potential by combining a physical and empirical approach. The
619 incorporation of upstream drainage area represents the supply of sediment and the surface
620 reflectance represents the different sediment characteristics of the land surface types (Parajuli
621 & Zender, 2017). The SSM dust potential scheme is a novel attempt to provide a global

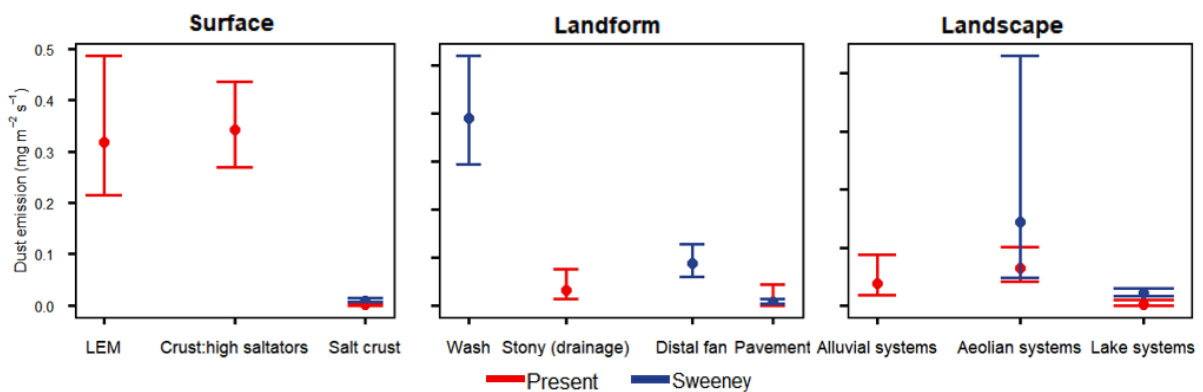
622 representation of erodibility at the landscape-scale, elegantly tuned to a maximum potential
623 represented by the Bodélé Depression in Chad. However, the landform-scale assessment of the
624 SSM presented here highlights that there are potential shortcomings in this erodibility map.
625 The dust source points identified with Landsat analysis indicates that most of the dust hot spots
626 in this area are situated in the terminal stages of the rivers as they near the Atlantic Ocean and
627 the coastal sabkhas (Figure 1, and see von Holdt et al., 2017). The SSM classification however
628 identifies areas with high emission potential significantly upstream of the confirmed dust
629 sources, including areas covering large areas of stony systems adjacent to alluvial and dry lake
630 sources (Figure 1e). Furthermore, rivers that are not significant dust sources, such as the
631 Swakop River (marked W in Figure 1 e) are identified as highly emissive in the SSM. The
632 relative absence of dust emission from the Swakop River is probably due to the incised nature
633 of this river combined with less topographic channelling of the high magnitude north-easterly
634 Bergwind compared to other more emissive rivers such as the Kuiseb River. In addition, many
635 parts of the Namib ephemeral rivers hold lush vegetation sustained by groundwater so that
636 significant vegetative roughness make sediments unavailable for entrainment (von Holdt &
637 Eckardt, 2018). As a result, the raised emission potential associated with enhanced alluvial
638 sediment supply is likely to be overestimated in the SSM. It follows that, in addition to the
639 preferential dust source areas identified by the classifications as applied to the Namib, the
640 influences of vegetation and topographic channelling would need to be adequately
641 parameterised in any dust model operating at this regional scale.

642 The assessment of SSM values for dust emission source points identified with Landsat and
643 measured for dust flux with the PI-SWERL indicate that the SSM scheme does not always
644 agree with the dust emission analysis presented here. The mean SSM value for the location of
645 all Landsat dust emission source points of 0.245 is just under half the maximum emission value
646 of 0.519 for the Namib region. Only 4.7% of the 2289 dust source points were classified in the
647 most emissive category >0.4 . Although the dust source points identified by Landsat do not
648 provide a continuous numerical quantification of the dust emission potential, this point source
649 inventory does identify areas which should be assigned values of high emissivity similar to the
650 method used by Parajuli & Zender (2017) in assigning a maximum value of 1 to the Bodélé
651 Depression. In addition, there is no clear relationship between the SSM emission values and
652 PI-SWERL emission results (Figure 5). Sites exhibiting high emission rates were not
653 necessarily classified as highly emissive according to the SSM. The most emissive site as
654 measured with the PI-SWERL (geometric mean dust flux: $0.2191 \text{ mg m}^{-2} \text{ s}^{-1}$) is situated on the

655 alluvial system terraces and have an SSM value of 0.25 and classified as stabilised sand deposit
656 by LSM. The highest SSM value (0.46) associated with the PI-SWERL runs was situated
657 within the stony system with a geometric mean dust flux of $0.0127 \text{ mg m}^{-2} \text{ s}^{-1}$, but classified as
658 lake system by the LSM due to the high surface reflectance of the quartz stone pavement . In
659 this context, the PI-SWERL provides a quantification of dust emission by which to compare
660 the dust emission potential from different surface types or landform- and landscape-scale
661 geomorphic units and offers a means to validate dust-emission schemes. Dust emission is
662 highly variable as indicated by the PI-SWERL dust flux measurements from the different
663 surfaces (Figure 3 Table 2). The small-scale variability that exists at sub-landform scale has
664 been seen to exert a clear effect on dust emission, so adequate representation of this variability
665 remains an important yet persistently challenging research goal. The combination of the
666 Landsat dust point source and PI-SWERL dust flux measurements at landform and sub-
667 landform scale can make contributions as input and validation for landscape-scale dust-
668 emission schemes.

669 The PI-SWERL potentially provides a standardised quantification of surface dust emissions
670 across different dust-producing regions, however comparison between different studies and
671 regions would require consistency in measured parameters and landform categories. We offer
672 an attempt at such a comparison by relating our results to Sweeney et al. (2011) from the
673 Mojave Desert, USA (referred to from hereon as SW2011), which shows agreement between
674 some of the measured surfaces, landforms and landscapes (Figure 7). An important legitimacy
675 to this comparison is that SW2011 tested at the same friction velocity ($u_* = 0.56 \text{ m s}^{-1}$) as our
676 study. The pavements in the stony systems and lakes systems compare well, whereas the
677 aeolian systems have good agreement between the geometric means and lower confidence
678 interval, but SW2011 report a much larger upper confidence interval. This is potentially due to
679 SW2011 consisting of many more replicates for this landform category (30 versus 3 for the
680 present study) and the dunes from the Mojave study being potentially situated adjacent to a
681 large pan. In turn, the dunes tested as part of the Namib study were situated near an ephemeral
682 river (Figure 1 a, approximately 1km south of Kuiseb delta), which would introduce potentially
683 finer material than dunes situated farther from such a source. This comparison illustrates the
684 importance of site-specific controls in accounting for the degree of inherent emission
685 variability for a given surface type and indicates their importance for understanding dust fluxes
686 as quantified from field testing in different locations. Aeolian systems can cover large areas
687 and their emission potential can vary greatly depending on factors such as dune type,

688 mineralogy and age (Bristow & Moller, 2018; Bullard et al., 2011). Dunes rarely produce
 689 distinct point sources visible on satellite imagery (as is the case for this Namib dataset, Figure
 690 2) but have the potential to be contributing sources of dust, albeit at low volumes, due to their
 691 areal extent. The dust emission potential of the Sand Seas of the Namib requires further
 692 investigation. Furthermore, what was classified as a stony system with drainage in this study
 693 probably most closely corresponds with a wash as per SW2011, which SW2011 determined to
 694 be considerably more emissive (SW2011 Wash geometric mean: $0.3915 \text{ mg m}^{-2} \text{ s}^{-1}$ vs
 695 geometric mean of stony system with drainage in present study: $0.0102 \text{ mg m}^{-2} \text{ s}^{-1}$). An
 696 important difference may well be down to fact that the stony systems with drainage classified
 697 as a landform in the Namib study typically did not feature a supply of sand available for
 698 saltation at the sites we tested, thereby they corresponded to a gravel-covered surface, rather
 699 than an LEM-dominated one (Figure 3). The LEM surface category from this study and the
 700 wash (Landform) from SW2011 were the two most emissive categories and represent a
 701 maximum emission value for the two studies. Despite LEM representing a surface type and
 702 wash a landform type, the upper limit of emission shows good agreement and again illustrates
 703 the importance of landform and surface interpretation.



704

705 **Figure 7 Comparison between the present study (red lines) and results from Sweeney et al.**
 706 **(2011) (blue lines) for the Mojave Desert, USA (both studies tested at $u_* = 0.56 \text{ m s}^{-1}$). Selected**
 707 **surfaces, landforms and landscape classes from both studies were chosen for comparison. The**
 708 **stone pavement and dry lakes (pans) show good agreement. Correspondence is shown between**
 709 **the two studies for the categories with maximum emission values (wash for Sweeney et al., 2011**
 710 **and LEM surface type for present study).**

711

712 The PI-SWERL measurements also can be used to assess sufficiency of sample size for
713 different scales of analysis, both for the current study and as a guide for future efforts. Because
714 our approach aggregates individual measurements with increasing scale of analysis, sample
715 sizes are reduced (from $n=128$ for individual PI-SWERL to $n=12$ for landscape systems, Table
716 1). This, in turn, affects the confidence intervals for each scale of analysis, with results
717 indicating that the current sampling effort may be insufficient to resolve significant differences
718 for some categories at the landform and landscape levels (Figure S4). Similarly, power analyses
719 demonstrate that larger sample sizes may be needed than the current effort for all scales of
720 analysis (Figure S5); requisite sample sizes are 60 per surface type, 22 transects per landform,
721 and 12 landforms per landscape. However, analysis of confidence intervals (which
722 demonstrates the impact of sample variance and sample size on uncertainty of the mean) are
723 generally preferred to a post hoc power analysis (Hoening & Heisey, 2001; Goodman & Berlin,
724 1994). Regardless, the above results indicate that, despite the extensive sampling effort
725 involved in our study (128 measurements), larger sample sizes may be needed in future work,
726 particularly at landform and landscape scales due to their large confidence intervals (Figure
727 S4).

728 Another potential approach to develop and improve dust-emission schemes with high-
729 resolution/fine-scale PI-SWERL data is to assess the factors that control the erodibility of the
730 surfaces. This is especially useful for factors that are represented in datasets that are available
731 globally, such as particle size and moisture data. The BRT analysis highlighted the significant
732 variables for PM_{10} dust flux measured with the PI-SWERL at a set friction velocity. The partial
733 dependency plots additionally highlight information on critical thresholds where dust emission
734 takes place or ceases on different surfaces (Figure 6). An important consideration for such
735 analysis is the choice of variables and the method of measurement. For example, the need to
736 sample moisture close to the surface meant that moisture was determined only as a relative
737 estimate at our test sites, not a quantified measure throughout the soil column. A more accurate
738 reflection of the moisture content would have been obtained by a gravimetric determination of
739 the top 1 cm of the soil surface, which would provide an indication of the soil moisture (%)
740 and its influence on the erosion threshold (u_{*r}). Furthermore, our approach of sampling soil
741 moisture in the upper 2 cm of the surface, rather than directly at the surface may have reduced
742 the performance of the BRT, particularly for samples taken from lake system playas and
743 sabkhas. When those data points are excluded, a critical moisture threshold of 2% is obtained
744 for the tested friction velocity of approximately 0.58 m s^{-1} . This moisture content threshold is

745 similar to the value of 0.02 g g^{-1} soil moisture obtained by Munkhtsetseg et al. (2016) above
746 which they observed that dust emission became significantly depressed.

747 Of further critical importance here are the environmental conditions at the time of surface
748 testing. The long-term Landsat record established that the ephemeral lakes are highly emissive
749 (Figure 2), in accordance with many dust sources (e.g. Gill, 1996; Reynolds et al., 2009;
750 Bullard et al., 2008; Ginoux et al., 2012), but their emissivity was relatively low during the PI-
751 SWERL testing due to the prevailing elevated humid conditions in proximity to the coast at the
752 time of PI-SWERL testing, as well as the hygroscopic saline surfaces and periodic shallow
753 water nature which can create wet playas (Reynolds et al., 2009; Sweeney et al., 2016). This
754 also raises an important issue regarding dust emission not captured with remote sensing. The
755 ephemeral lakes of the Namib are highly emissive during the Bergwind events that coincide
756 with the overpass of the polar-orbiting satellites (MODIS and Landsat). However, the
757 conditions are dominated by high relative humidity along the coast which also prevailed at the
758 time of testing. The effect of such environmental controls is underscored by the fact that during
759 high relative humidity along the coast, MODIS and Landsat indicate the alluvial flood terraces
760 remained emissive, whereas the coastal-adjacent ephemeral lakes were non-emissive. As a
761 result, we could be over-estimating the emissions from these lakes. Relative humidity should
762 be a standard measurement that has to be recorded at the time of PI-SWERL testing. The BRT
763 analysis should also be extended by testing at different friction velocities to determine the
764 threshold at which emission is initiated. Furthermore, by identifying the significant variables,
765 a set of surface characterisation tests can be developed that should be included when measuring
766 emission potential. Combining a standard set of surface characterisation tests and dust flux
767 measurements from different well-known hot spots around the world can be used to
768 substantially improve dust-emission schemes.

769 **6 Conclusion**

770 This study provides a ground-based assessment of recently proposed dust-emission mapping
771 schemes that highlights limitations in our ability to represent dust emission potential at large
772 scales. The novel combination of a high resolution (Landsat-derived), sub-landscape scale
773 inventory of actively eroding parts of the Namib Desert together with the ground-based
774 measurement of dust emission rates using PI-SWERL at known point sources allows a
775 qualitative and quantitative evaluation of two approaches to classify emission potential. Our
776 findings demonstrate that point measurements of emission, coupled with characterization of

777 surface properties (soil moisture, degree of crusting, particle size and elemental composition
778 as proxy for mineralogy) at the time of testing, can provide valuable information for assessing
779 and potentially improving larger-scale schemes for predicting dust emission potential.

780 The combination of a physical and empirical approach such as that used to create the Sediment
781 Supply Map (SSM) (Parajuli & Zender, 2017) is useful at capturing the relevant factors for
782 surface erodibility and emission potential based on globally-applicable data sets. The SSM
783 represents sediment supply on the basis of drainage area (a proxy for hydrology) and surface
784 reflectance, but in the Namib, alluvial systems, seen here to be areas of potentially high
785 emission, are mapped by SSM as lying significantly upstream and over a greater extent
786 compared to the distribution of active dust source points identified via remote sensing. In the
787 case of emission potential determined from user-defined geomorphic mapping such as that
788 offered by PDS (Bullard et al., 2011), challenges include determining the inputs and effort
789 required to apply the scheme across a region, as well as variability of emission within a given
790 class. Our sub-landform measurements indicate that variability of emission rates remains an
791 inherent problem for each of the emission classification schemes examined in our study.
792 However, results such as ours can be used to further parameterize the range of emission fluxes
793 for land surface classes in dust-cycle models. The current study reveals how PI-SWERL dust-
794 emission measurements provide a relative quantification of landform emissivity which can
795 provide modelers with the range of emission fluxes for given geomorphic classes in emission
796 potential mapping schemes.

797 Use of a Boosted Regression Tree (BRT) model identifies significant surface characteristics
798 and critical thresholds related to dust emission that can be used to inform dust models. The
799 BRT analysis for the Namib Desert highlighted the importance of soil moisture content, crust
800 strength and particle size kurtosis, with critical thresholds for dust emission additionally
801 dependent on gravel density and the presence of sand and silt. Our approach provides a
802 framework for obtaining site-specific values in other dust-source regions and may help to
803 standardize datasets for global dust emission modeling. A standardised set of surface
804 characterisation tests combined with dust flux measurements would offer regional and global
805 datasets of relative emission potential and thereby provide utility for developing dust emission
806 schemes toward improved dust emission modeling.

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820 **References**

- 821 Ashpole, I. & Washington, R., 2013. A new high-resolution central and western Saharan
822 summertime dust source map from automated satellite dust plume tracking. *Journal of*
823 *Geophysical Research: Atmospheres*, 118(13), pp.6981-6995.
- 824 Bacon, S.N., McDonald, E. V., Amit, R., Enzel, Y. & Crouvi, O. 2011. Total suspended
825 particulate matter emissions at high friction velocities from desert landforms. *Journal*
826 *of Geophysical Research: Earth Surface*. 116(3):1–17. DOI: 10.1029/2011JF001965.
- 827 Baddock, M.C., Bullard, J.E. & Bryant, R.G., 2009. Dust source identification using MODIS:
828 a comparison of techniques applied to the Lake Eyre Basin, Australia. *Remote Sensing*
829 *of Environment*, 113(7), pp.1511-1528.
- 830 Baddock, M.C., Ginoux, P., Bullard, J.E. and Gill, T.E. (2016), Do MODIS-defined dust
831 sources have a geomorphological signature? *Geophysical Research Letters*.
832 43(6):2606–2613. DOI: 10.1002/2015GL067327.
- 833 Belnap, J. & Gillette, D.A., 1998. Vulnerability of desert biological soil crusts to wind
834 erosion: the influences of crust development, soil texture, and disturbance. *Journal of*
835 *Arid Environments*, 39(2), pp.133-142.
- 836 Blott, S.J. & Pye, K., 2001. GRADISTAT: a grain size distribution and statistics package for
837 the analysis of unconsolidated sediments. *Earth Surface Processes and*
838 *Landforms*, 26(11), pp.1237-1248.
- 839 Bristow, C.S. and Moller, T.H., 2018. Dust Production by Abrasion of Eolian Basalt Sands:
840 Analogue for Martian Dust. *Journal of Geophysical Research: Planets*, 123(10),
841 pp.2713-2731.
- 842 Bryant, R.G., Bigg, G.R., Mahowald, N.M., Eckardt, F.D. & Ross, S.G. 2007. Dust emission
843 response to climate in southern Africa. *Journal of Geophysical Research*. 112(D9):1–
844 17. DOI: 10.1029/2005JD007025.
- 845 Buck, B.J., King, J. and Etyemezian, V., 2011. Effects of salt mineralogy on dust emissions,
846 Salton Sea, California. *Soil Science Society of America Journal*, 75(5), 1971-1985.
- 847 Bullard, J.E., Harrison, S.P., Baddock, M.C., Drake, N., Gill, T.E., McTainsh, G. & Sun, Y.,
848 2011. Preferential dust sources: A geomorphological classification designed for use in

849 global dust-cycle models. *Journal of Geophysical Research: Earth Surface*, 116(F4).
850 DOI: 10.1029/2011JF002061

851 Bullard, J., Baddock, M., McTainsh, G. & Leys, J. 2008. Sub-basin scale dust source
852 geomorphology detected using MODIS. *Geophys. Res. Lett.*. 35:1–6. DOI:
853 10.1029/2008GL033928.

854 Bullard, J.E., Mctainsh, G.H. & Pudmenzky, C. 2004. Aeolian abrasion and modes of fine
855 particle production from natural red dune sands : an experimental study.
856 *Sedimentology*. 51:1103–1125. DOI: 10.1111/j.1365-3091.2004.00662.x.

857 Cornelis, W.M., Gabriels, D., Hartmann, R., 2004. A conceptual model to predict the
858 deflation threshold shear velocity as affected by near-surface soil water: II.
859 Calibration and validation. *Soil Science Society of America Journal* 68, 1162–1168.

860 Crouvi, O., Amit, R., Enzel, Y., Porat, N. & Sandler, A. 2008. Sand dunes as a major
861 proximal dust source for late Pleistocene loess in the Negev Desert , Israel.
862 *Quaternary Research*. 70:275–282. DOI: 10.1016/j.yqres.2008.04.011.

863 Dansie, A.P., Thomas, D.S.G., Wiggs, G.F.S. and Munkittrick, K.R. 2018 [Spatial variability of](#)
864 [ocean fertilising nutrients in the dust-emitting ephemeral river catchments of](#)
865 [Namibia](#). *Earth surface processes and landforms*, 43: 563-578.

866 Eckardt, F.D. and Kuring, N. 2005. SeaWiFS identifies dust sources in the Namib Desert.
867 *International journal of remote sensing*, 26(19), pp.4159-4167.

868 Eitel, B., Blümel, W.D., Hüser, K. and Mauz, B., 2001. Dust and loessic alluvial deposits in
869 Northwestern Namibia (Damaraland, Kaokoveld): sedimentology and palaeoclimatic
870 evidence based on luminescence data. *Quaternary International*, 76, pp.57-65.

871 Elith, J., Leathwick, J.R. & Hastie, T. 2008. A working guide to boosted regression trees.
872 *Journal of Animal Ecology*. 77(4):802–813. DOI: 10.1111/j.1365-2656.2008.01390.x.

873 Etyemezian, V., Nikolich, G., Ahonen, S., Pitchford, M., Sweeney, M., Purcell, R., Gillies, J.
874 & Kuhns, H. 2007. The Portable In Situ Wind Erosion Laboratory (PI-SWERL): A
875 new method to measure PM10 windblown dust properties and potential for emissions.
876 *Atmospheric Environment*. 41(18):3789–3796. DOI:
877 10.1016/j.atmosenv.2007.01.018.

- 878 Gill, T. E. (1996), Eolian sediments generated by anthropogenic disturbance of playas:
879 Human impacts on the geomorphic system and geomorphic impacts on the human
880 system, *Geomorphology*, 17, 207–228, doi:10.1016/0169-555X(95)00104-D.
- 881 Gillette, D.A. Adams, J., Muhs, D. & Kihl, R., 1982. Threshold friction velocities and rupture
882 moduli for crusted desert soils for the input of soil particles into the air. *Journal of*
883 *Geophysical Research*, 87(C11), p 9003. Available at:
884 <http://doi.wiley.com/10.1029/JC087iC11p09003>.
- 885 Gillette, D. A., 1999 A qualitative geophysical explanation for “hot spot” dust emitting
886 source regions, *Contrib. Atmos. Phys.*, **72**, 67– 77.
- 887 Gillies, J.A., Nickling, W.G. and King, J., 2006. Aeolian sediment transport through large
888 patches of roughness in the atmospheric inertial sublayer. *Journal of Geophysical*
889 *Research: Earth Surface*, 111(F2).
- 890 Gillies, J. A. (2013), Fundamentals of aeolian sediment transport: Dust emissions and
891 transport—Near surface, in *Treatise on Geomorphology*, edited by J. Shroder and N.
892 Lancaster, pp. 43–63, Academic Press, San Diego, Calif.
- 893 Ginoux, P., Chin, M., Tegen, I., Prospero, J.M., Holben, B., Dubovik, O. and Lin, S.J., 2001.
894 Sources and distributions of dust aerosols simulated with the GOCART
895 model. *Journal of Geophysical Research: Atmospheres*, 106(D17), pp.20255-20273.
- 896 Ginoux, P., Prospero, J.M., Gill, T.E., Hsu, N.C. & Zhao, M. 2012. Global-Scale Attribution
897 of Anthropogenic and Natural Dust Sources and Their Emission Rates Based on
898 Modis Deep Blue Aerosol Products. 1–36. DOI: 10.1029/2012RG000388.
- 899 Goodman, S.N. and Berlin, J.A., 1994. The use of predicted confidence intervals when
900 planning experiments and the misuse of power when interpreting results. *Annals of*
901 *internal medicine*, 121(3), pp.200-206.
- 902 Goossens, D. & Buck, B., 2009. Dust emission by off-road driving: Experiments on 17 arid
903 soil types, Nevada, USA. *Geomorphology*, 107(3), pp.118-138.
- 904 Goudie, A. and Viles, H., 2015. Climate. In *Landscapes and Landforms of Namibia* (pp. 37-
905 46). Springer Netherlands.

- 906 Grini, A., Myhre, G., Zender, C.S. & Isaksen, I.S., 2005. Model simulations of dust sources
907 and transport in the global atmosphere: Effects of soil erodibility and wind speed
908 variability. *Journal of Geophysical Research: Atmospheres*, 110(D2).
- 909 Haustein, K., Washington, R., King, J., Wiggs, G., Thomas, D. & Menut, L., 2015. Testing
910 the performance of state-of-the-art dust-emission schemes using DO4Models field
911 data. *Geosci Model Dev Discuss*, 7, pp.5739-5789.
- 912 Hijmans, R.J., Phillips, S., Leathwick, J. and Elith, J., 2016. dismo: species distribution
913 modeling.—R package ver. 1.0-15.
- 914 Hoenig, J.M. and Heisey, D.M., 2001. The abuse of power: the pervasive fallacy of power
915 calculations for data analysis. *The American Statistician*, 55(1), pp.19-24.
- 916 Hsu, N.C., Tsay, S.C., King, M.D. and Herman, J.R., 2004. Aerosol properties over bright-
917 reflecting source regions. *IEEE Transactions on Geoscience and Remote
918 Sensing*, 42(3), pp.557-569.
- 919 Huang, J., Ge, J. & Weng, F. 2007. Detection of Asia dust storms using multisensor satellite
920 measurements. *Remote Sensing of Environment*. 110(2):186–191. DOI:
921 10.1016/j.rse.2007.02.022.
- 922 Ishizuka, M., Mikami, M., Yamada, Y., Zeng, F. and Gao, W., 2005. An observational study
923 of soil moisture effects on wind erosion at a gobi site in the Taklimakan Desert.
924 *Journal of Geophysical Research: Atmospheres*, 110(D18).
- 925 Jacobson, P. J., Jacobson, K. N., & Seely, M. K. 1995. *Ephemeral rivers and their
926 catchments: sustaining people and development in western Namibia*. Desert Research
927 Foundation of Namibia.
- 928 King, J., Etyemezian, V., Sweeney, M., Buck, B.J. & Nikolich, G. 2011. Dust emission
929 variability at the Salton Sea, California, USA. *Aeolian Research*. 3(1):67–79. DOI:
930 10.1016/j.aeolia.2011.03.005.
- 931 Lee, J. A., Gill, T. E., Mulligan, K. R., Dominguez Acosta, M., Perez, A. E., 2009. Land
932 use/land cover and point sources of the 15 December 2003 dust storm in southwestern
933 North America. *Geomorphology*, 105(December 2003), 18–27.

- 934 Lee, J. A., Baddock, M.C., Mbuh, M.J. & Gill, T.E. 2012. Geomorphic and land cover
935 characteristics of aeolian dust sources in West Texas and eastern New Mexico, USA.
936 *Aeolian Research*. 3(4):459–466. DOI: 10.1016/j.aeolia.2011.08.001.
- 937 Mahowald, N., Luo, C., Corral, J. & Zender, C.S. 2003. Interannual variability in
938 atmospheric mineral aerosols from a 22-year model simulation and observational data.
939 *Journal of Geophysical research*. 108(D12). DOI: 10.1029/2002JD002821.
- 940 Mahowald, N.M. and Dufresne, J.L., 2004. Sensitivity of TOMS aerosol index to boundary
941 layer height: Implications for detection of mineral aerosol sources. *Geophysical*
942 *Research Letters*, 31(3).
- 943 Marticorena, B. and Bergametti, G., 1995. Modeling the atmospheric dust cycle: 1. Design of
944 a soil-derived dust emission scheme. *Journal of Geophysical Research:*
945 *Atmospheres*, 100(D8), pp.16415-16430.
- 946 McFadden, L.D., 2013. Strongly dust-influenced soils and what they tell us about landscape
947 dynamics in vegetated aridlands of the southwestern United States. *Geological Society*
948 *of America Special Papers*, 500, pp.501-532.
- 949 McKenna-Neuman, C. & Nickling, W.G., 1989. A theoretical and wind tunnel investigation
950 of the effect of capillary water on the entrainment of sediment by wind. *Canadian*
951 *Journal of Soil Science*, 69(1), pp.79-96.
- 952 Neuman, C.M. and Maxwell, C., 2002. Temporal aspects of the abrasion of microphytic
953 crusts under grain impact. *Earth Surface Processes and Landforms: The Journal of*
954 *the British Geomorphological Research Group*, 27(8), pp.891-908.
- 955
- 956 Munkhtsetseg, E., Shinoda, M., Gillies, J.A., Kimura, R., King, J. & Nikolich, G. 2016.
957 Relationships between soil moisture and dust emissions in a bare sandy soil of
958 Mongolia. *Particuology*. 28:131–137. DOI: 10.1016/j.partic.2016.03.001.
- 959 Okin, G.S. and Gillette, D.A., 2001. Distribution of vegetation in wind-dominated
960 landscapes: Implications for wind erosion modeling and landscape processes. *Journal*
961 *of Geophysical Research: Atmospheres*, 106(D9), pp.9673-9683.

962 O’Loingsigh, T., R.M. Mitchell, S.K. Campbell, N.A. Drake, G.H. McTainsh, N.J. Tapper,
963 and D.L. Dunkerley (2015), Correction of dust event frequency from MODIS Quick-
964 Look imagery using in-situ aerosol measurements over the Lake Eyre Basin,
965 Australia, *Rem. Sens. Environ.*, 169, 222-231. doi: 10.1016/j.rse.2015.08.010.

966 Paradis, E. and Schliep, K., 2018. ape 5.0: an environment for modern phylogenetics and
967 evolutionary analyses in R. *Bioinformatics*, 35(3), pp.526-528.

968 Parajuli, S.P. & Zender, C.S. 2017. Connecting geomorphology to dust emission through
969 high-resolution mapping of global land cover and sediment supply. *Aeolian Research*.
970 27:47–65. DOI: 10.1016/j.aeolia.2017.06.002.

971 Parajuli, S.P., Yang, Z. & Kocurek, G. 2014. Mapping erodibility in dust source regions
972 based on geomorphology, meteorology, and remote sensing. *Journal of Geophysical*
973 *Research: Earth Surface*, 119, DOI: 10.1002/2014JF003095.

974 Pinheiro J, Bates D, DebRoy S, Sarkar D, R Core Team (2018). *nlme: Linear and Nonlinear*
975 *Mixed Effects Models*. R package version 3.1-137, <URL: [https://CRAN.R-projec](https://CRAN.R-project.org/package=nlme)
976 [t.org/package=nlme](https://CRAN.R-project.org/package=nlme)>.

977 Prospero, J.M., Ginoux, P., Torres, O., Nicholson, S.E. & Gill, T.E., 2002. Environmental
978 characterization of global sources of atmospheric soil dust identified with the Nimbus
979 7 Total Ozone Mapping Spectrometer (TOMS) absorbing aerosol product. *Reviews of*
980 *geophysics*, 40(1).

981 Pye, K. & Tsoar, H., 1990. *Aeolian Sand and Sand Dunes*. Unwin Hyman, London. 396p.

982 QGIS Development Team, 2016. QGIS Geographic Information System. Open Source
983 Geospatial Foundation. URL <http://qgis.osgeo.org>

984 R Core Team (2017). *R: A language and environment for statistical computing*. R Foundation
985 for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.

986

987 Raupach, M.R., Gillette, D.A. and Leys, J.F., 1993. The effect of roughness elements on wind
988 erosion threshold. *Journal of Geophysical Research: Atmospheres*, 98(D2), pp.3023-
989 3029.

990

- 991 Ravi, S., D'odorico, P., Breshears, D.D., Field, J.P., Goudie, A.S., Huxman, T.E., Li, J., Okin,
 992 G.S., Swap, R.J., Thomas, A.D. & Van Pelt, S., 2011. Aeolian processes and the
 993 biosphere. *Reviews of Geophysics*, 49(3)
- 994 Resane, T., Freiman, T. and Annegarn, H., 2004. The day of the white rain: origin of unusual
 995 dust deposition in Johannesburg, South Africa. *South African journal of*
 996 *science*, 100(9), pp.483-487.
- 997 Reynolds, R.L., Bogle, R., Vogel, J., Goldstein, H. & Yount, J. 2009. Dust emission at
 998 Franklin Lake Playa , Mojave Desert (USA): Response to meteorological and
 999 hydrologic changes 2005-2008. *Natural Resources and Environmental Issues*.
 1000 15(1):Article 18.
- 1001 Sankey, J.B., Glenn, N.F., Germino, M.J., Gironella, A.I.N. and Thackray, G.D. (2010),
 1002 Relationships of aeolian erosion and deposition with LiDAR-derived landscape
 1003 surface roughness following wildfire. *Geomorphology*, 119(1-2), 135-145.
- 1004 Sankey, J.B., Eitel, J.U., Glenn, N.F., Germino, M.J. and Vierling, L.A., 2011. Quantifying
 1005 relationships of burning, roughness, and potential dust emission with laser altimetry of
 1006 soil surfaces at submeter scales. *Geomorphology*, 135(1-2), pp.181-190.
- 1007 Schepanski, K., Tegen, I., Laurent, B., Heinold, B. & Macke, A., 2007. A new Saharan dust
 1008 source activation frequency map derived from MSG-SEVIRI IR-channels. *Geophys.*
 1009 *Res. Lett.*, 34(18).
- 1010 Schepanski, K., Tegen, I. and Macke, A., 2012. Comparison of satellite based observations of
 1011 Saharan dust source areas. *Remote Sensing of Environment*. 123:90–97. DOI:
 1012 10.1016/j.rse.2012.03.019.
- 1013 Shao, Y.P., Raupach, M.R. and Leys, J.F., 1996. A model for predicting aeolian sand drift
 1014 and dust entrainment on scales from paddock to region. *Soil Research*, 34(3), 309-
 1015 342.
- 1016 Shao, Y., Wyrwoll, K.H., Chappell, A., Huang, J., Lin, Z., McTainsh, G.H., Mikami, M.,
 1017 Tanaka, T.Y., Wang, X. and Yoon, S., 2011. Dust cycle: An emerging core theme in
 1018 Earth system science. *Aeolian Research*, 2(4), 181-204.
- 1019 Shi, M., Yang, Z.L., Stenchikov, G.L., Parajuli, S.P., Tao, W. and Kalenderski, S., 2016.
 1020 Quantifying the impacts of landscape heterogeneity and model resolution on dust

- 1021 emissions in the Arabian Peninsula. *Environmental modelling & software*, 78, pp.106-
1022 119.
- 1023 Strong C.L., J.E. Bullard, G.H. McTainsh, C. Dubois, and M.C. Baddock (2010), Impact of
1024 wildfire on interdune ecology and sediments: an example from the Simpson Desert,
1025 Australia, *J. Arid Env.*, 74: 1577-1581. doi:10.1016/j.jaridenv.2010.05.032.
- 1026 Sweeney, M., Etyemezian, V., Macpherson, T., Nickling, W., Gillies, J., Nikolich, G., and
1027 McDonald, E. (2008), Comparison of PI-SWERL with dust emission measurements
1028 from a straight-line field wind tunnel. *Journal of Geophysical Research: Earth*
1029 *Surface*, 113(1), 1–12. doi: 10.1029/2007JF000830
- 1030 Sweeney, M.R., McDonald, E. V. and Etyemezian, V. (2011), Quantifying dust emissions
1031 from desert landforms, eastern Mojave Desert, USA, *Geomorphology*. 135: 21–34.
1032 DOI: 10.1016/j.geomorph.2011.07.022.
- 1033 Sweeney, M.R., McDonald, E.V. and Markley, C.E., (2013), Alluvial sediment or playas:
1034 what is the dominant source of sand and silt in desert soil vesicular A horizons,
1035 southwest USA. *Journal of Geophysical Research: Earth Surface*, 118(1), 257-275.
- 1036 Sweeney, M.R., Zlotnik, V.A., Joeckel, R.M. and Stout, J.E. (2016), Geomorphic and
1037 hydrologic controls of dust emissions during drought from Yellow Lake playa, West
1038 Texas, USA. DOI: 10.1016/j.jaridenv.2016.05.007.
- 1039 Taramelli, A., Pasqui, M., Barbour, J., Kirschbaum, D., Bottai, L., Busillo, C., Calastrini, F.,
1040 Guarnieri, F. and Small, C., 2013. Spatial and temporal dust source variability in
1041 northern China identified using advanced remote sensing analysis. *Earth Surface*
1042 *Processes and Landforms*, 38(8), pp.793-809.
- 1043 Thomas, D.S., Durcan, J.A., Dansie, A. and Wiggs, G.F., 2017. Holocene fluvial valley fill
1044 sources of atmospheric mineral dust in the Skeleton Coast, Namibia. *Earth Surface*
1045 *Processes and Landforms*, 42(12), pp.1884-1894.
- 1046 Vickery, K.J., Eckardt, F.D. & Bryant, R.G. 2013. A sub-basin scale dust plume source
1047 frequency inventory for southern Africa, 2005-2008. *Geophysical Research Letters*.
1048 40(19):5274–5279. DOI: 10.1002/grl.50968.
- 1049 Vickery, K. J. & Eckardt, F. D., 2013. Dust emission controls on the lower Kuiseb River
1050 valley, Central Namib. *Aeolian Research*, 10, 125–133.

- 1051 von Holdt, J.R., Eckardt, F.D. & Wiggs, G.F.S. 2017. Landsat identifies aeolian dust
1052 emission dynamics at the landform scale. *Remote Sensing of Environment*. 198:229–
1053 243. DOI: 10.1016/j.rse.2017.06.010.
- 1054 von Holdt, J.R. and Eckardt, F.D., 2018. Dust activity and surface sediment characteristics of
1055 the dustiest river in southern Africa: the Kuiseb River, Central Namib. *South African*
1056 *Geographical Journal*, 100(1), pp.104-121.
- 1057 Wang, X., Lang, L., Hua, T., Wang, H., Zhang, C. & Wang, Z. 2012. Characteristics of the
1058 Gobi desert and their significance for dust emissions in the Ala Shan Plateau (Central
1059 Asia): An experimental study. *Journal of Arid Environments*. 81:35–46. DOI:
1060 10.1016/j.jaridenv.2012.01.014.
- 1061 Washington, R., Todd, M., Middleton, N.J. & Goudie, A.S. 2003. Dust-storm source areas
1062 determined by the total ozone monitoring spectrometer and surface observations.
1063 *Annals of the Association of American Geographers*. 93(2):297–313. DOI:
1064 10.1111/1467-8306.9302003.
- 1065 Webb, N.P. & Strong, C.L. 2011. Soil erodibility dynamics and its representation for wind
1066 erosion and dust emission models. *Aeolian Research*. 3(2):165–179.
- 1067 Wiggs, G.F.S., Baird, A.J. and Atherton, R.J., 2004. The dynamic effects of moisture on the
1068 entrainment and transport of sand by wind. *Geomorphology*, 59(1-4), pp.13-30.
- 1069 Zender, C.S., Bian, H. & Newman, D., 2003. Mineral Dust Entrainment and Deposition
1070 (DEAD) model: Description and 1990s dust climatology. *Journal of Geophysical*
1071 *Research: Atmospheres*, 108(D14).
- 1072