

1 Mesophilic anaerobic co-digestion of the sewage sludge with glycerine:

2 Effect of solid retention time.

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11

ABBREVIATION LIST

AD: anaerobic digestion

COD: chemical oxygen demand

GP: gas production

HRT: Hydraulic retention time

MP: methane production

OLR: organic loading rate

SMP: specific methane production

SRT. Solid retention time

TS: total solids

VFA: volatile fatty acids

VS: volatile solids

WWTP: wastewater treatment plant

12 **Abstract**

13 The main objective of this paper was to examine the effect of the increase of organic
14 loading rates (OLRs) (by reducing solid retention time, SRT, from 20 d to 5 d) in
15 single-phase mesophilic anaerobic co-digestion of the sewage sludge with glycerine
16 (1% v/v). Experimentally, it was confirmed that anaerobic co-digestion of these
17 biowastes in steady-state conditions can achieve $85\pm 5\%$ of volatile fatty acid (VFA)
18 reduction at SRTs between 20 and 9 d, with a methane production around $0.8 \text{ l CH}_4/\text{l/d}$.
19 Decreases in the SRT not only allow maintaining the sludge stability and the biogas
20 production, but it also implies an increase in the waste that could be treated and lower
21 cost. Therefore, mesophilic anaerobic co-digestion of sewage sludge and glycerin at
22 SRT lower than 20 d is possible and preferable, due to it is more economical and
23 environmental friendly.

24

25 **Keywords:** anaerobic co-digestion; biogas, SRT, mesophilic, sewage sludge, glycerine

26

27 **1. Introduction**

28 Disposal of sewage sludge as a main byproduct generated in wastewater treatment
29 plants (WWTPs) is a major challenge which typically represents up to 50% of the
30 overall operating costs of a WWTP [1,2]. Anaerobic sludge digestion as a reliable
31 technology employed worldwide to stabilize organics and reduce solids, destroy
32 pathogens and produce biogas as the source of energy [2–7]. Due to the advantages of
33 anaerobic digestion (AD), many research studies have sought to optimize the AD of
34 sludge, including the interesting option of the co-digestion process [8], which increases
35 the load of biodegradable organic matter and produces a higher biogas yield. Recent
36 studies have been demonstrated the efficacy of anaerobic co-digestion of municipal
37 sludge or solid waste together with readily biodegradable organic substances, such as
38 glycerol, a major by-product of biodiesel production [1,9–14]. Production of 100 kg of
39 biodiesel yields approximately 10 kg of glycerin waste as a co-product. Numerous
40 industries such as pharmaceutical, cosmetics and food processing, use refined glycerol
41 as a raw input material. However, the glycerol generated, as a co-product of biodiesel
42 production requires purification before being suitable for use in these industries.
43 Therefore, the glycerol is often considered waste stream instead of a co-product [11],
44 which makes its disposal a fundamental environmental concern. Most of the recent
45 studies about AD of glycerin and sludge or municipal solid waste to improve the
46 methane production (MP) have been focused to identify the optimal concentration (%
47 v/v) of glycerin that have to be added into the substrate [1,9–11,13,14] and very few
48 have studied the effect on semi-continuous or continuous feeding regime [1,11] and
49 none at optimization of solid retention time (SRT) during the anaerobic co-digestion of
50 sludge and glycerol in a single phase reactor to MP. SRT optimization is important,

51 since low SRT are preferred for real application, reducing the volume of the anaerobic
52 digester and the WWTP cost. Based on these premises the present study has been
53 developed. The experimental protocol was designed to examine the effect of the
54 increase of organic loading rate (by reducing the solid retention time, SRT, from 20 d to
55 5 d) on the efficiency of stirred tank reactor treating sewage sludge and glycerin and to
56 report on its steady-state performance. The experimental reactor was subjected to a
57 program of steady-state operation over a range of solid retention times, SRTs, from 20 -
58 5 d and organic loading rates (OLRs), from 1.03 to up to 4.05 g COD/l/d in order to
59 evaluate its treatment capacity.

60

61 **2. Materials and Methods**

62 **2.1. Substrates and inoculum**

63 Experimental work was carried out with sewage sludge samples (mixed primary sludge
64 and activated sludge) from Cadiz-San Fernando WWTP (located in Cadiz-Spain, which
65 handles more than 50,000 m³ of wastewater daily). All the sludge samples were
66 characterized on reception at the laboratory and were kept under refrigeration at 4 °C
67 before they were used for the experiments so as to prevent biodegradation. This sludge
68 was mixed with 1% v/v glycerol commercial household Panreac, which constituted the
69 reactor feed. According to Fountoulakis et al. [14], the most appropriate concentration
70 of glycerol in co-digestion with sewage sludge in anaerobic processes is 1%. The main
71 characteristics of the sewage sludge are summarized in Table 1.

72 Regarding to the inoculum, it was collected from the mesophilic anaerobic digester
73 present at the same WWTP. The pH, total solids (TS) and volatile solids (VS) were 7.5±
74 0.2; 32.0 ± 2.0 g TS/kg and 18.0 ± 0.2 g VS/kg, respectively.

75

76 **2.2. Experimental equipment and operation conditions**

77 The laboratory-scale reactor used for this study operates in a semi-continuous stirred
78 tank reactor and in the mesophilic range (35°C). The equipment consists of a reactor
79 with a stainless steel vessel that is agitated and heated and that has a total volume of 5 L
80 and a working volume of 4.5 L (Figure 1). No biomass recycling was used; the
81 hydraulic retention time (HRT) and the Solid Retention Time (SRT) are equal. The
82 reactor features a lid that allows it to be sealed to maintain anaerobic conditions within
83 the reactor.

84 The stainless steel lid has three openings (one for the biogas outlet, a feed inlet and
85 another opening for the stirring system). The bottom of the reactor has a release valve
86 used for sampling the material inside the reactor, which is made possible by the sealing
87 system between the vessel and the cap. The assembly includes an agitator that achieves
88 the homogenisation of waste using stainless steel blade scrapers. To maintain the
89 operating temperature (mesophilic, 35 °C), the reactor is heated by recirculating water
90 through a thermostatic jacket. Biogas is collected in 40-L Tedlar bags and a special
91 syringe is used for sampling gases.

92 Regarding to the operation conditions, the reactor was fed with sewage sludge and
93 glycerin (1%) without nutrients and pH correction once a day (semi-continuous regime).

94 Based on information found in the literature and the previous experience of the group
95 [15,16], SRTs of 20, 15, 9, 7 and 5 days were selected for study until the process
96 breakdown. Figure 2 describe the SRTs and its corresponding OLRs applied to the
97 reactor during the experiment. The overall duration of the experiments was 255 d.

98

99 **2.3. Analytical methods**

100 To characterise the waste and the inoculum, as well as to monitor the effluent of the
101 process, the following were analysed: pH, alkalinity, volatile fatty acids (VFA), total
102 chemical oxygen demand (TCOD), total solids (TS) and volatile solids (SV). These
103 analyses were conducted in accordance with standard methods (APHA, 1995) and
104 Zahedi et al. [9]. The gas volume produced in the reactors was directly measured using
105 a high-precision flow gas meter: Ritter_ drum-type gas meter TG-01-Series (Wet-Type).
106 VFA were determined by gas chromatography, using a gas chromatograph (Shimadzu
107 GC-2010) equipped with a flame ionization detector (FID) and a capillary column filled
108 with Nukol. The gas volume produced in the reactor was directly measured using a
109 high-precision flow gas meter: TG-01-Series (Wet-Type) Ritter drum-type gas meter.
110 The biogas composition was determined by gas chromatography separation
111 (SHIMADZU GC-2010). H₂, CH₄, CO₂, O₂ and N₂ were analysed by means of a
112 thermal conductivity detector (TCD) employing a Supelco Carboxen 1010 Plot column.
113 Samples were taken using a 1 ml Dynatech Gastight gas syringe under the following
114 operating conditions: split = 100; constant pressure in the injection port (70 kPa); 2 min
115 at 40 °C; ramped at 40 °C/min until 200 °C; 1.5 min at 200 °C; detector temperature: 250
116 °C; and injector temperature: 200°C. Helium was used as carrier gas (266.2 ml/min).
117 Commercial mixtures of H₂, CH₄, CO₂, O₂, N₂ and H₂S (Abelló Linde S.A.) were used
118 to calibrate the system.
119 Gas volume and composition were measured daily; in the effluent, the pH was
120 measured daily in all condition assayed. VS, TCOD, alkalinity and VFA were analysed
121 approximately three times a week.

122

123 **3. Results and discussion**

124 This section discusses the evolution of the main variables during the semi-continuous
125 mesophilic anaerobic digestion process, such as pH, VFA, alkalinity, TCOD, VS and
126 biogas production and composition. All the values correspond to the analytical
127 determinations in steady conditions (except at 5 d SRT, because destabilization was
128 observed). The number of determinations in steady conditions considered to present the
129 averages and standard values of biogas and pH were 35, 20, 12, 11 and 9 for 20, 15, 10,
130 9 and 7 d SRT, respectively; and the analyses (in steady conditions) considered to
131 present the averages and standard values VFA, COD, alkalinity and VS were 9, 8, 12, 9
132 and 8 for 20, 15, 10, 9 and 7 d SRT. Data of 5 d SRT are unstable and they only are
133 represented to show the system destabilization.

134 **3.1. Process stability: pH evolution and VFA/alkalinity ratio**

135 The stability of the process was evaluated based on pH and the VFA/alkalinity ratio at
136 different SRT tested [9,18]. pH was used to evaluate the stability along the process and
137 VFA/alkalinity ratio was used to establish under which conditions (SRT) the mesophilic
138 anaerobic co-digestion of the sewage sludge with glycerine could operate without
139 jeopardising its safety.

140 pH is a basic parameter for control of the anaerobic degradation process. Figure 3 shows
141 the evolution of pH during the semi-continuous mesophilic study. In this figure, vertical
142 lines are included to indicate the changes in SRT and red horizontal dashed line
143 indicates the 7.0 pH. Initially, with an SRT of 20 d, 15 d and 12 d, pH values stabilise at
144 approximately 7.3-7.8, the optimum pH for the activity of methanogenic
145 microorganisms [9,19]. With an SRT of 9 d, the pH decrease during the first 5 d until it
146 reached a value 6.44, as result of the increased OLR feed, but finally it was increase and

147 stabilised at 7.3-7.5, without the addition of an external agent (the reactor was no
148 feeding per two days and it was recovered). The initial decrease in pH when an increase
149 in the OLR is applied into the reactor may be due to the initial imbalance between the
150 metabolic activities of microbial groups. When the added load is increased, the
151 acidogenic microorganisms respond quickly, given their high specific growth rate and
152 generate more VFA. However, methanogenic archaea are slower and require more time
153 to grow and reach the population size necessary to degrade the excess of VFA.
154 Finally, the decreases in the SRT at 5 d supposed the pH dropped to values below 6 and
155 the reactor was not recovered. At this condition, the pH decreases as a result of the
156 accumulation of VFA in the reactor due to methanogenic archaea were not able to
157 degrade the excess of VFA produced by hydrolytic-acidogenic bacteria [15,16,20],
158 indicating acidification of the medium and thus destabilize the process. In short, taking
159 to account the pH values it could be said that single-phase mesophilic anaerobic co-
160 digestion of the sewage sludge with glycerine (1% v/v) is totally steady at SRTs
161 between 20 and 7 d.

162 As previously mentioned, VFA/alkalinity ratio was also considered, to establish under
163 which conditions (SRT) the mesophilic anaerobic co-digestion of the sewage sludge
164 with glycerine could operate without jeopardising its safety. This parameter have been
165 used to evaluate the stability of the process during the AD of waste and glycerol [9].
166 The medium values of these ratios are shown Figure 4. Values between 0.1 and 0.4
167 (equiv. acetic acid/equiv. CaCO_3) indicate favourable operating conditions without the
168 risk of acidification. In this figure, horizontal dashed line indicates the 0.4 value. At
169 SRT higher than 7, this parameter was under 0.4 (equiv. acetic acid/equiv. CaCO_3)
170 indicating a proper performance. At 7 d SRT, VFA/alkalinity ratio was slightly higher

171 than the optimum values, indicating risk of acidification. Therefore, at 7 d of SRT while
172 the pH values were maintained high so as to allow methanogenic activity, the acids
173 generated during the acidogenic are not totally consumed and some of them are
174 accumulated in the system, thus start to affect the activity of the anaerobic consortia and
175 a reduction in the organic matter removal is observed, as will be explained later. At but
176 at SRT of 5 d this parameter was too high, indicating total system destabilization. This
177 effect (high values for both, pH and VFA/alkalinity ratio) has been also detected in
178 under non-stable AD process of glycerin and biowastes [9]. In short, taking to account
179 the pH and VFA/alkalinity ratios values it could be said that single-phase mesophilic
180 anaerobic co-digestion of the sewage sludge with glycerine (1% v/v) is totally steady at
181 SRTs between 20 and 9 d. At 7 d of SRT the acids generated during the acidogenic
182 phase start to accumulate in the system. Therefore, although in this study the system has
183 been able to maintain the pH values around 7.0-7.4, 7 d of SRT could be considered a
184 critical time to operate to mesophilic anaerobic co-digestion of the sewage sludge with
185 glycerine (1% v/v), especially in real-industrial WWTP digester.

186

187 **3.2. Organic matter removal**

188 Figure 5 shows the removal efficiencies of VFA, TCOD and VS (as %) in the
189 mesophilic reactor for different SRTs. For 20 d SRT, VFA, TCOD reaches
190 approximately 60% TCOD removal, 50% VS removal and 85% VFA removal. Similar
191 values are found in the other stages of the operation until 7 d SRT. At 7 d a huge
192 decrease in the organic matter was detected consumption was detected VFA, TCOD and
193 VS reaches a small value (25% VFA removal; 15% TCOD removal and 30% VS
194 removal). At 5 d the organic matters removals were lower that between 4-10%. Taking

195 into account these parameters we can ensure that at 7 d STR the reactor is not degrading
196 properly organic matter and at 5 d SRT the reactor is not able to assimilate the ORL
197 feed and consequently the pH decreases in the system and produces an accumulation of
198 VFA, and it implied an increase in VFA/alkalinity ratio, as seen in previous sections. In
199 short, breakdown efficiency starts at 7 d SRT and retention times shorter than 7 d are
200 insufficient for a stable digestion of mixed sludge and glycerin (1% v/v).

201 Logically, the best results for TCOD concentration and VFA, in terms of the quality of
202 the effluent, were obtained in the range 20-9 d SRT. The total acidity, expressed as the
203 total amount VFA represented by acetic acid and TCOD, exhibits stable values in the
204 effluent from the mesophilic reactor in the range 395 - 155 mg acetic/l and 6-8 g O₂/l,
205 respectively, at SRTs between 20 and 9 d. When the SRT is changed to 7 d, a
206 significant difference in average total acidity is observed, with a value of 1640 mg
207 acetic/l, due to the increased organic load supplied to the system. This trend illustrates
208 the initial destabilisation caused by the reduction in the SRT, as discussed above.

209 However, at the end of the 5 d SRT, the average acidity values are close to 12000 mg
210 acetic/l of acetic acid indicating total destabilization in the system. The increase in the
211 VFA when glycerine is added into the feed has been reported in several studies
212 [9,11,14,21]. Holm-Nielsen et al. [21] studied the anaerobic digestion of a mixture of
213 manure, waste from food industries and glycerol added to the reactor gradually. The
214 authors observed the accumulation of volatile fatty acids and glycerol in the reactor with
215 the addition of 3.5 to 6.5% of glycerol (v/v), from the 16th to the 19th day of the
216 experiment, which caused the inhibition of the methanogenic phase. Razaviarani et al.
217 [11] observed that the accelerated increase in VFA concentration in the test digester and
218 decrease in the biogas CH₄ content suggest that methanogens inhibition occurred at

219 supplementation of glycerin of 1.8% (v/v) were added to municipal sludge at SRT of 20
220 d. In a similar study, Fountoulakis et al. [14] reported that adding 3% (v/v) glycerine to
221 sewage sludge resulted in VFA accumulation and process instability at 23–25 d of SRT.
222 Zahedi et al. [9] study explores the effect of five different glycerol supplementations
223 (0%, 0.1%, 0.25%, 0.5% and 1%) on effluent characteristics, anaerobic consortia and
224 MP in batch mode and they observed that during the acidogenic phase of anaerobic
225 VFA were accumulated at supplementation of glycerin of 0.5% (v/v). However the
226 effect of tat different SRT and for a constant value of glycerin, as the present study is
227 worked, has not been related yet. The destabilization in the present paper is not due to
228 the high values of glycerin as the other researchers discussed above, since at AD of
229 sewage sludge and glycerin (1% v/v) at SRT between of 20-9 d VFA were very low
230 (395 - 155 mg acetic/l). The destabilization was due to over load produced by a decrease
231 in the SRT.

232 **3.3. Biogas**

233 Fig. 6 shows the medium values of biogas production (l /l reactor/d) and methane
234 production (l CH₄ /l reactor/d) at every condition tested from SRT of 20 d to 5 d. At
235 SRT from 20 to 9 d SRT, the medium values of GP and MP were ranged between 1.21-
236 1.43 l/d and between 0.6-0.9 l CH₄/l/d respectively. For SRT lower than 9 d, the
237 tendency changes and a drop in both GP and MP are observed indicating overload or
238 destabilization, especially at 5 d SRT in which an extremely decrease was noted,
239 coinciding with the total destabilization of the system. It was in line with the other
240 parameters (pH and organic matter removal decrease and VFA increase). For the higher
241 SRT (20d, 15 d, 12 d and 9 d), the values of the MP and specific methane production
242 (SMP, ml of methane per TCOD consumed) were ranged between 0.6-0.9 l CH₄/l/d and

243 between 0.28-0.33 l CH₄/ g TCOD, respectively and these results are in line with others
244 about AD of sewage sludge and glycerine [1,11,12].

245 **3.4 Optimal conditions of the mesophilic AD process of sewage sludge and glycerin**

246 AD process of sewage sludge and glycerin for stabilizing sludge and for obtaining
247 renewable energy was carried out at six different SRTs (from 20 d to 5 d) or six OLRs
248 (from 1.03 g to 4.05 g COD/l/d) (Figure 2). As, the GP, MP and SMP were more or less
249 constant (except to 7 d and 5 d), the quality of the effluent and the adaptive capacity
250 were the parameters selected to determine the optimum operating condition for AD
251 process of glycerin and sewage sludge from Cadiz-San Fernando WWTP.

252 Considering the pH values, the low values of VFA, the high organic matter removal it
253 could be said that AD process of sewage sludge and glycerin at SRT of 15 d and 12 d
254 can be as effective as SRT of 20d. 9 d of SRT seems to be very low because it is close
255 to the SRT in which VFA start increasing and organic matter removal start decreasing
256 (SRT of 7 d). Therefore digesters could operate at 15-12 d instead of 20 d without
257 jeopardising its safety, since at supplementations of 1% (v/v) glycerin is stable up to 7 d
258 SRT. It means a reduction in the reactor cost (initial cost) of the AD up to 25-40%,
259 compared to AD process of sewage sludge and glycerin at SRT of 20 d. In addition, the
260 operational cost of anaerobic co-digestion will be reduce (lower time to heat and mix
261 the waste).

262 In short AD process of sewage sludge and glycerin at SRT lower than 20 d is possible
263 and preferable, due to it is more economical and environmental friendly. Decreases in
264 the SRT not only allow maintaining the sludge stability and the biogas production, but it
265 also implies an increase in the waste that could be treated and lower initial (lower

266 volume of reactor) operational cost (lower volume/time to heat and mix) in a real
267 process. It is an important fact, due to sludge management is a serious issue since up to
268 one-half of the costs of operating WWTPs is associated with sludge treatment and
269 disposal [2,7,22] and it has been estimated that 4 billion gallons of crude glycerol will
270 be produced each year by the biodiesel market reached [9,23]. Therefore every process
271 to allow treat more waste in a shorter time or produce a reduction in the cost (initial cost
272 or operating cost) of the WWTP have to be highlighted.

273

274 **Conclusion**

275 The effectiveness of the glycerin supplementation during the AD of sewage sludge at
276 different SRT was assessed in this study. The following conclusions have been
277 obtained:

- 278 • Co-digestion of glycerin and sludge is totally stable at SRT between 20 and 9 d.
279 No significant differences in methane production and organic matter removal
280 were detected under these conditions.
- 281 • AD process of sewage sludge and glycerin at SRT lower than 20 d is possible
282 and preferable, due to it is more economical and environmental friendly. In a
283 real WWTP operate at 15-12 d instead of 20 d could suppose a reduction in the
284 WWTP cost of the AD of biowaste.

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293

294 **References**

- 295 [1] Razaviarani V, Buchanan ID. Anaerobic co-digestion of biodiesel waste glycerin
296 with municipal wastewater sludge: Microbial community structure dynamics and
297 reactor performance. *Bioresour Technol* 2015;182:8–17.
298 doi:10.1016/j.biortech.2015.01.095.
- 299 [2] Zahedi S, Icaran P, Yuan Z, Pijuan M. Assessment of free nitrous acid pre-
300 treatment on a mixture of primary sludge and waste activated sludge: Effect of
301 exposure time and concentration. *Bioresour Technol* 2016;216:870–5.
302 doi:10.1016/j.biortech.2016.06.038.
- 303 [3] Riau V, De la Rubia MA, Pérez M. Assessment of solid retention time of a
304 temperature phased anaerobic digestion system on performance and final sludge
305 characteristics. *J Chem Technol Biotechnol* 2012;87:1074–82.
306 doi:10.1002/jctb.3709.
- 307 [4] Forster-Carneiro T, Riau V, Pérez M. Mesophilic anaerobic digestion of sewage
308 sludge to obtain class B biosolids: Microbiological methods development.
309 *Biomass and Bioenergy* 2010;34:1805–12. doi:10.1016/j.biombioe.2010.07.010.
- 310 [5] Liao X, Li H, Zhang Y, Liu C, Chen Q. Accelerated high-solids anaerobic
311 digestion of sewage sludge using low-temperature thermal pretreatment. *Int*
312 *Biodeterior Biodegradation* 2016;106:141–9. doi:10.1016/j.ibiod.2015.10.023.
- 313 [6] Wu L-J, Higashimori A, Qin Y, Hojo T, Kubota K, Li Y-Y. Upgrading of
314 mesophilic anaerobic digestion of waste activated sludge by thermophilic pre-
315 fermentation and recycle: Process performance and microbial community
316 analysis. *Fuel* 2016;169:7–14. doi:10.1016/j.fuel.2015.11.091.

- 317 [7] Peces M, Astals S, Clarke WP, Jensen PD. Semi-aerobic fermentation as a novel
318 pre-treatment to obtain VFA and increase methane yield from primary sludge.
319 *Bioresour Technol* 2016;200:631–8. doi:10.1016/j.biortech.2015.10.085.
- 320 [8] Mata-Alvarez J, Dosta J, Macé S, Astals S. Codigestion of solid wastes: a review
321 of its uses and perspectives including modeling. *Crit Rev Biotechnol*
322 2011;31:99–111. doi:10.3109/07388551.2010.525496.
- 323 [9] Zahedi S, Solera R, García-Morales JL, Ennouri H, Sales D. Evaluation of the
324 effect of glycerol supplementation on the anaerobic digestion of real municipal
325 solid waste in batch mode. *Fuel* 2017;193:15–21. doi:10.1016/j.fuel.2016.12.024.
- 326 [10] Zahedi S, Solera R, García-Morales JL, Sales D. Effect of the addition of
327 glycerol on hydrogen production from industrial municipal solid waste. *Fuel*
328 2016;180:343–7. doi:10.1016/j.fuel.2016.04.063.
- 329 [11] Razaviarani V, Buchanan ID, Malik S, Katalambula H. Pilot scale anaerobic co-
330 digestion of municipal wastewater sludge with biodiesel waste glycerin.
331 *Bioresour Technol* 2013;133:206–12. doi:10.1016/j.biortech.2013.01.101.
- 332 [12] Rivero M, Solera R, Perez M. Anaerobic mesophilic co-digestion of sewage
333 sludge with glycerol: Enhanced biohydrogen production. *Int J Hydrogen Energy*
334 2014;39:2481–8. doi:10.1016/j.ijhydene.2013.12.006.
- 335 [13] Fountoulakis MS, Manios T. Enhanced methane and hydrogen production from
336 municipal solid waste and agro-industrial by-products co-digested with crude
337 glycerol. *Bioresour Technol* 2009;100:3043–7.
338 doi:10.1016/j.biortech.2009.01.016.

- 339 [14] Fountoulakis MS, Petousi I, Manios T. Co-digestion of sewage sludge with
340 glycerol to boost biogas production. *Waste Manag* 2010;30:1849–53.
341 doi:10.1016/j.wasman.2010.04.011.
- 342 [15] Fernández Rodríguez J, Pérez M, Romero LI. Mesophilic anaerobic digestion of
343 the organic fraction of municipal solid waste: Optimisation of the semicontinuous
344 process. *Chem Eng J* 2012;193-194:10–5. doi:10.1016/j.cej.2012.04.018.
- 345 [16] Montañés R, Pérez M, Solera R. Anaerobic mesophilic co-digestion of sewage
346 sludge and sugar beet pulp lixiviation in batch reactors: Effect of pH control.
347 *Chem Eng J* 2014;255:492–9. doi:10.1016/j.cej.2014.06.074.
- 348 [17] APHA. *Am Public Heal Assoc Stand Methods Exam Water Wastewater* 1995.
- 349 [18] Rincón B, Borja R, González JM, Portillo MC, Sáiz-Jiménez C. Influence of
350 organic loading rate and hydraulic retention time on the performance, stability
351 and microbial communities of one-stage anaerobic digestion of two-phase olive
352 mill solid residue. *Biochem Eng J* 2008;40:253–61.
353 doi:10.1016/j.bej.2007.12.019.
- 354 [19] Dahunsi SO, Oranusi S, Owolabi JB, Efeovbokhan VE. Synergy of Siam weed
355 (*Chromolaena odorata*) and poultry manure for energy generation: Effects of pre-
356 treatment methods, modeling and process optimization. *Bioresour Technol*
357 2016;225:409–17. doi:10.1016/j.biortech.2016.11.123.
- 358 [20] Zahedi S, Sales D, Romero LI, Solera R. Optimisation of the two-phase dry-
359 thermophilic anaerobic digestion process of sulphate-containing municipal solid
360 waste: Population dynamics. *Bioresour Technol* 2013;148:443–52.

- 361 [21] Holm-Nielsen, Lomborg C., Oleskowicz-Popiel P, Esbensen K. On- line near
362 infrared monitoring of glycerol-boosted anaerobic digestion processes: evaluation
363 of process analytical technologies. *Biotechnol Bioeng* 2008;99:302–13.
- 364 [22] Lens P, Hamelers B, Hoitink H, Bidlingmaier W. *Resource Recovery and Reuse*
365 *in Organic Solid Waste Management*. IWA Publ London, UK 2004.
- 366 [23] Maru BT, López F, Kengen SWM, Constantí M, Medina F. Dark fermentative
367 hydrogen and ethanol production from biodiesel waste glycerol using a co-
368 culture of *Escherichia coli* and *Enterobacter* sp. *Fuel* 2016;186:375–84.
369 doi:10.1016/j.fuel.2016.08.043.

370

371

372 **Figure captions**

373 **Figure 1.** The laboratory-scale reactor used for this study.

374 **Figure 2.** Experimental conditions applied during the assay (SRTs and OLRs).

375 **Figure 3.** pH evolution along the process at different SRTs (● 20 d, ◆ 15 d, ▼ 12 d,

376 ▼ 9 d, □ 7 d, + 5 d) .

377 **Figure 4.** VFA/alkalinity ratio (equiv. acetic acid/equiv. CaCO₃) at different SRTs

378 tested.

379 **Figure 5.** Medium values of organic matter removal: VFA, TCOD and VS.

380 **Figure 6.** GP and MP at different SRTs tested.

381

382 **Table 1. Main characteristics of sewage sludge.**

Parameters	Sewage Sludge
pH	5.65±0.11
Conductivity (mS/cm)	9.88±1.25
TS (g/kg)	45.02±4.52
VS (g/kg)	34.59±5.05
TCOD (g O₂/l)	49.41±5.53
TOC (g/l)	15.83±2.36

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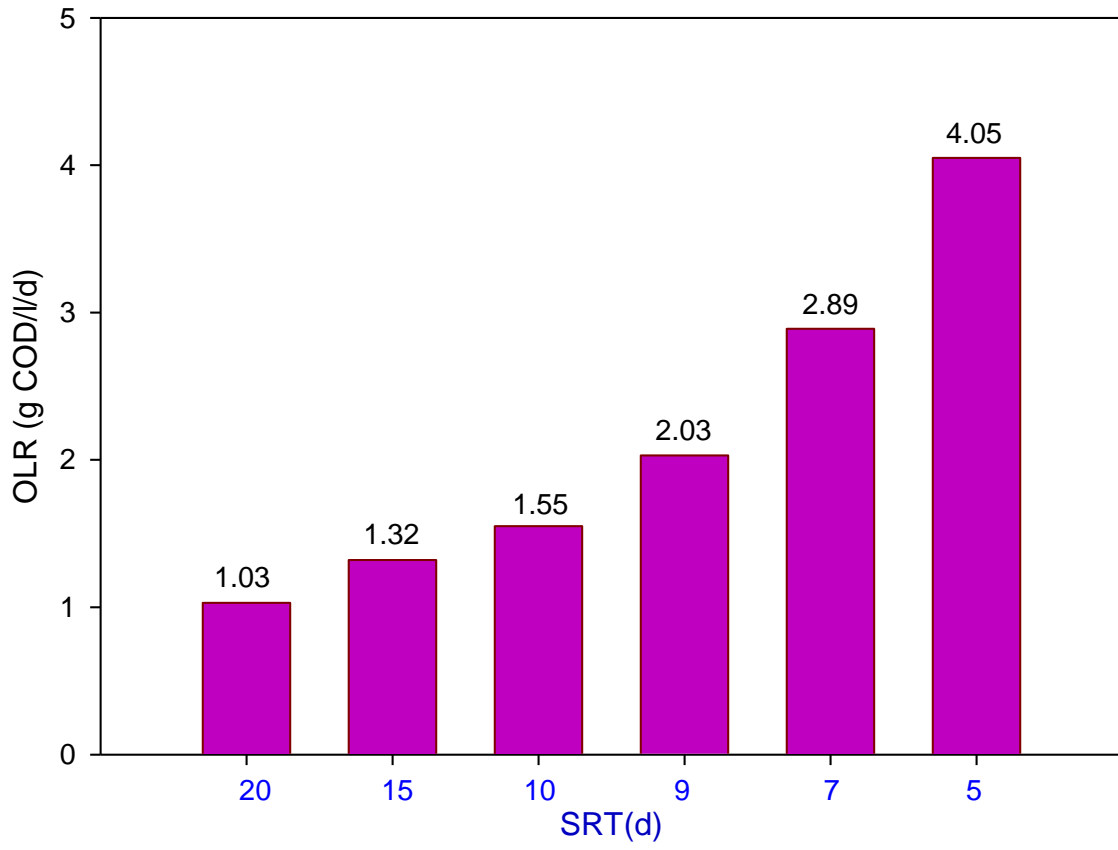
385 **Figure 1**



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388 **Figure 2**

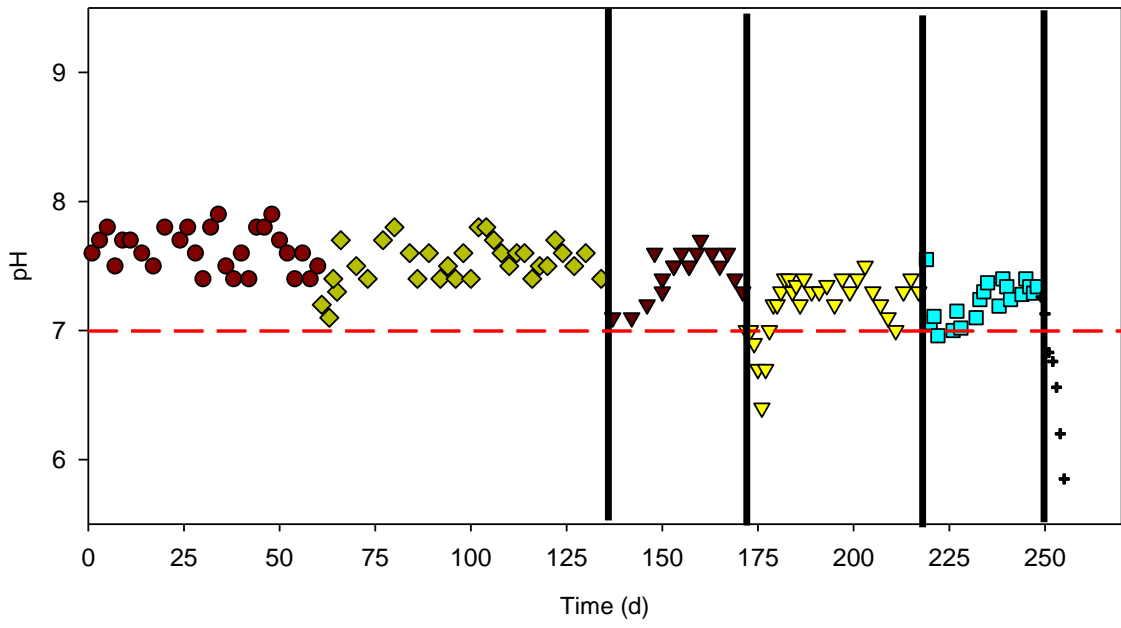


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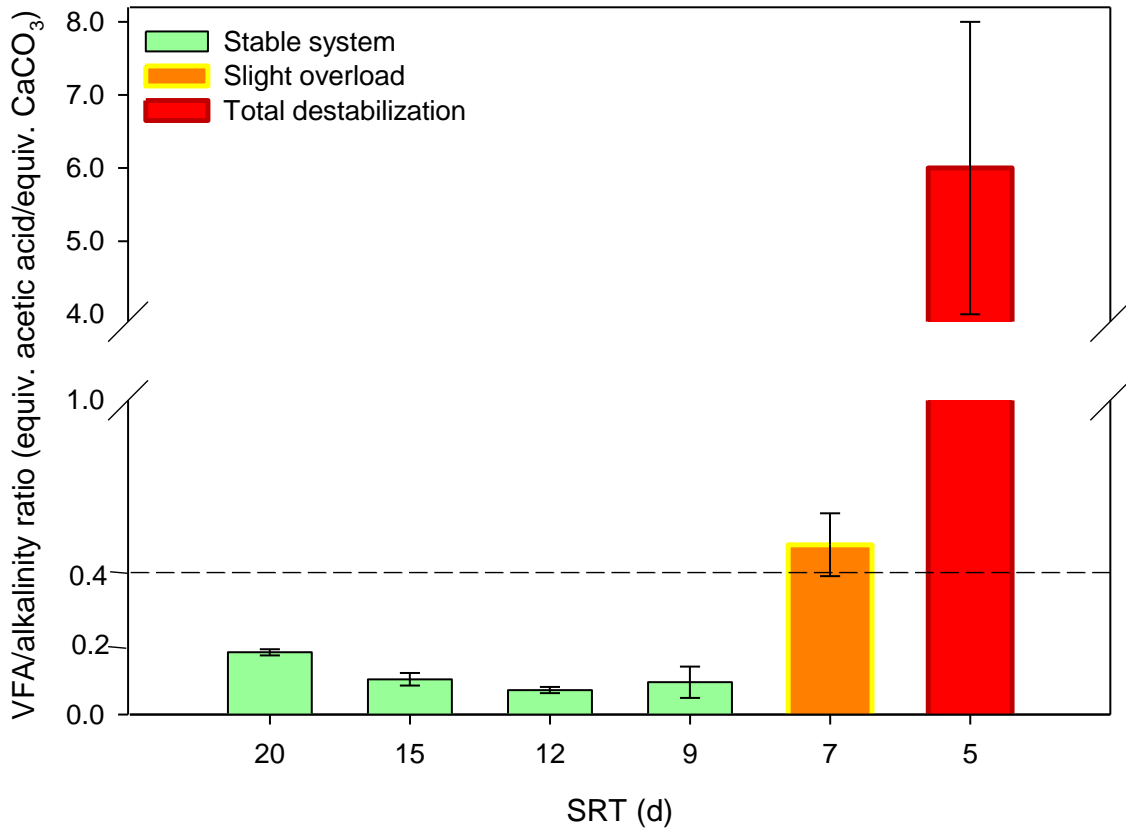
392 **Figure 3**



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395 **Figure 4**

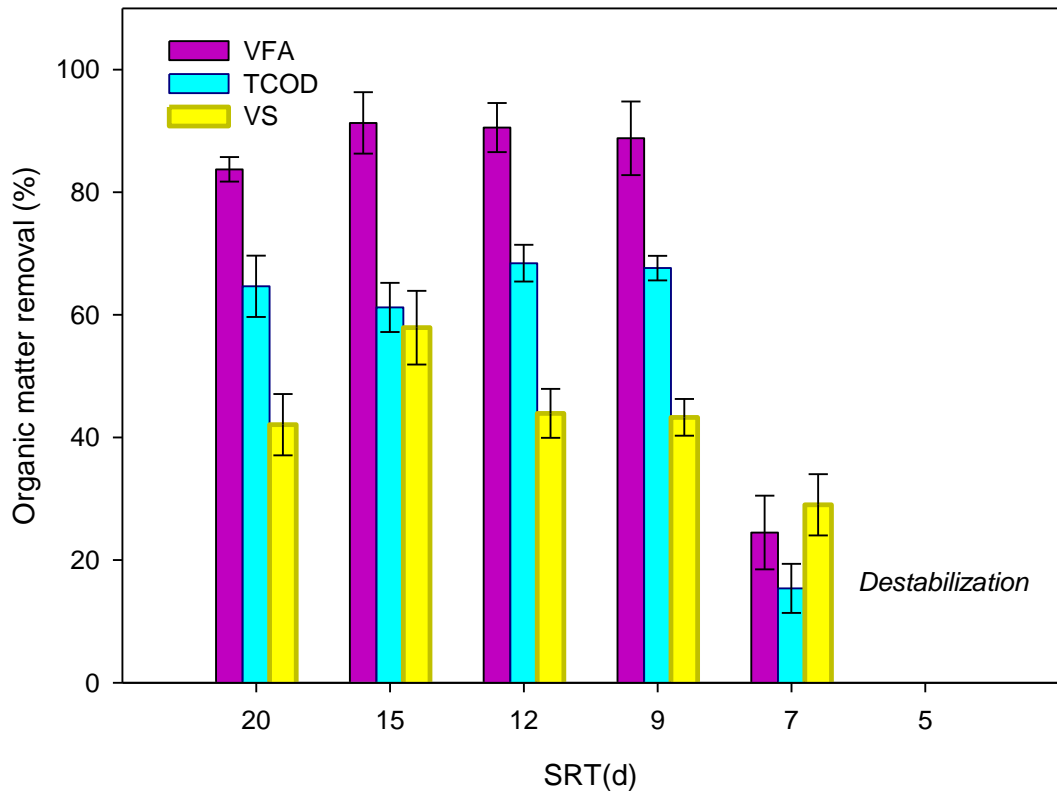


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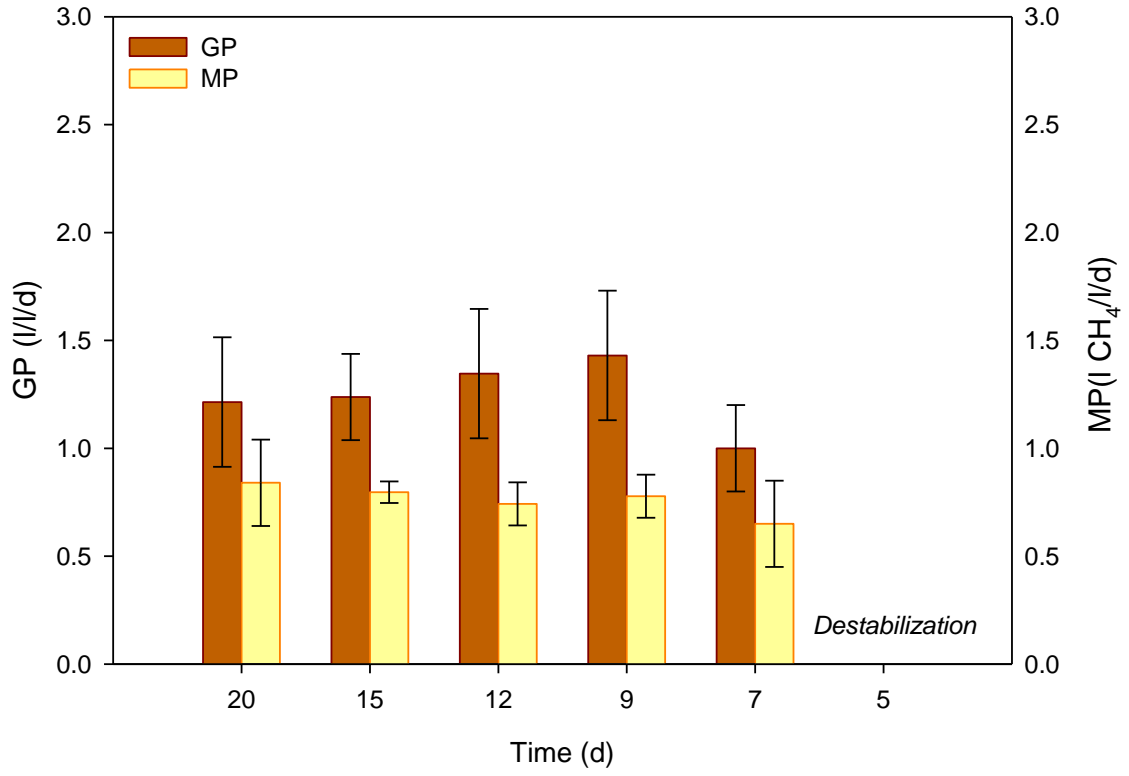
399 **Figure 5**



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402 **Figure 6**



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