

1 **TITLE**

2 **Assessing the physiological responses of the gastropod *Crepidula fornicata* to**
3 **predicted ocean acidification and warming**

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19 **Running title:** Responses of *C. fornicata* to OA and warming

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22

23

24 **Abstract**

25 Organisms inhabiting coastal waters naturally experience diel and seasonal physico-
26 chemical variations. According to various assumptions, coastal species are either considered to
27 be highly tolerant to environmental changes or, conversely, living at the thresholds of their
28 physiological performance. Therefore, these species are either more resistant or more sensitive,
29 respectively, to ocean acidification and warming. Here, we focused on *Crepidula fornicata*, an
30 invasive gastropod that colonized bays and estuaries on northwestern European coasts during
31 the 20th century. Small (< 3 cm in length) and large (> 4.5 cm in length), sexually mature
32 individuals of *C. fornicata* were raised for 6 months in three different $p\text{CO}_2$ conditions (390,
33 750 and 1400 μatm) at four successive temperature levels (10, 13, 16 and 19°C). At each
34 temperature level and in each $p\text{CO}_2$ condition, we assessed the physiological rates of
35 respiration, ammonia excretion, filtration and calcification on small and large individuals.
36 Results show that, in general, temperature positively influenced respiration, excretion and
37 filtration rates in both small and large individuals. Conversely, increasing $p\text{CO}_2$ negatively
38 affected calcification rates, leading to net dissolution in the most drastic $p\text{CO}_2$ condition (1400
39 μatm) but did not affect the other physiological rates. Overall, our results indicate that *C.*
40 *fornicata* can tolerate ocean acidification, particularly in the intermediate $p\text{CO}_2$ scenario.
41 Moreover, in this eurythermal species, moderate warming may play a buffering role in the
42 future responses of organisms to ocean acidification.

43

44 **Keywords:** calcification, coastal system, invasive species, metabolism, mollusk, $p\text{CO}_2$,
45 temperature

46

47

48

49 **Introduction**

50 Predictions indicate that coastal ecosystems will be strongly affected by ocean
51 acidification and warming, currently two of the most prominent anthropogenic processes
52 influencing marine life (Harley et al. 2006). Due to the increase in atmospheric CO₂ partial
53 pressure ($p\text{CO}_2$), pH in surface waters is predicted to decline by 0.06 to 0.32 units and sea
54 surface temperatures to increase by 1.0 to 3.7°C by the end of the century, depending on the
55 Intergovernmental Panel on Climate Change (IPCC) representative concentration pathway
56 considered (Stocker et al. 2013). Modifications in seawater carbonate chemistry due to ocean
57 acidification lead to a decrease in carbonate ion concentrations (CO_3^{2-}) (Orr et al. 2005) and a
58 reduction in the calcium carbonate saturation state (Ω), which regulates the thermodynamics of
59 calcium carbonate (CaCO_3) precipitation (Feely et al. 2009). In estuarine and coastal waters,
60 pH is more variable than in the open ocean due to intense biological and biogeochemical
61 processes (Andersson and Mackenzie 2011). In these habitats, ocean acidification and warming
62 will shift the baselines, exacerbate natural variations in pH and temperature, and probably
63 threaten the communities living there (Waldbusser and Salisbury 2013).

64 Mollusks constitute a major taxonomic group in estuarine and coastal waters in terms
65 of community structure and ecosystem functioning (Gutiérrez et al. 2003). Because most marine
66 mollusk taxa accumulate significant amounts of CaCO_3 to form protective external shells, they
67 may be sensitive to the changes in pH and carbonate chemistry induced by ocean acidification
68 (for review, see Gazeau et al. 2013), although recent studies have shown that some species
69 could be resistant to elevated $p\text{CO}_2$ (Range et al. 2011; Ries et al. 2009). Along with direct
70 impacts on calcification, high CO₂ concentrations may also have indirect effects on metabolism
71 by disturbing the extracellular acid-base equilibrium, leading to general internal acidosis
72 (Melzner et al. 2009). These potential shifts in acid-base homeostasis have the potential to
73 change organisms' energy balance (Pörtner et al. 2005).

74 In mollusks, the effects of elevated $p\text{CO}_2$ and/or decreased pH alone are highly species-
75 specific (see review in Gazeau et al. 2013), and depend on species sensitivity and any existing
76 compensation mechanisms (Michaelidis et al. 2005). To better estimate future ocean
77 acidification effects on mollusk species, various physiological processes have been studied in
78 bivalves and gastropods such as respiration (Beniash et al. 2010; Bibby et al. 2007), excretion
79 (Fernandez-Reiriz et al. 2011; Liu and He 2012), feeding (Fernandez-Reiriz et al. 2012;
80 Marchant et al. 2010), immune response (Bibby et al. 2008; Matozzo et al. 2012) and protein
81 or enzyme production (Matozzo et al. 2013; Tomanek et al. 2011). However, few studies have
82 simultaneously assessed the responses of more than three physiological processes to ocean
83 acidification and warming. The concomitant increase in seawater temperature and $p\text{CO}_2$ are
84 likely to affect mollusk metabolism because, in addition to changes in gas solubility and the
85 proportion of carbon species (Zeebe 2011), temperature also strongly affects physiological and
86 biochemical reactions (Cossins and Bowler 1987). Because warming can modulate the
87 metabolism responses to ocean acidification (Ivanina et al. 2013; Melatunan et al. 2013),
88 investigations of both pH and temperature effects are valuable for understanding the responses
89 of mollusks in the future ocean.

90 One of the most abundant and widespread shelled mollusks on the French northwestern
91 Atlantic and Channel coasts is the slipper limpet *Crepidula fornicata*, Linnaeus 1758
92 (Blanchard 1997). This gastropod native to the northeastern American coast was introduced in
93 Europe at the end of the 19th century, mainly via oysters imported for farming (Blanchard,
94 1995). It then colonized European coasts from southern Sweden to southern France, becoming
95 invasive in some places (Blanchard 1997). *C. fornicata* lives in shallow sites, especially in bays
96 and estuaries where it can reach very high densities of more than 1000 individuals per m^2
97 (Blanchard 1995). This species is known to be highly robust to environmental stress, in
98 particular temperature and salinity (Diederich and Pechenik 2013; Noisette et al. 2015),

99 parameters that have diel and seasonal variations in these coastal habitats. Established *C.*
100 *forficata* populations have largely affected biodiversity and ecosystem functioning in terms of
101 sediment modifications (Ehrhold et al. 1998), changes in faunal assemblages (De Montaudouin
102 et al. 1999) and trophic structure (Chauvaud et al. 2000). This species also affects benthic
103 biogeochemical cycles by enhancing filtration, metabolic activities, CaCO₃ production, and the
104 recycling of nutrients and dissolved carbon back into the pelagic ecosystem (Martin et al. 2006;
105 Martin et al. 2007; Ragueneau et al. 2002)

106 Although *C. forficata* is likely highly tolerant to environmental fluctuations, the
107 combined effects of decreased pH and increased temperature may push this species away from
108 its physiological optimum. Thus the objective of this work was to quantify the respiration,
109 ammonia excretion, filtration and calcification responses of small and large specimens of *C.*
110 *forficata* in different temperature and *p*CO₂ conditions. Investigating the physiology of this key
111 engineer in some coastal ecosystems in a context of climate change is one way to better
112 understand the sensitivity of this species and its potential future ecological impact.

113

114 **Methods**

115

116 *Sampling site and in situ conditions*

117 *C. forficata* stacks were collected by SCUBA divers on 30 November 2011, in Morlaix
118 Bay (northwestern Brittany, France), at the “Barre des Flots” site (3°53.015'W; 48°40.015'N)
119 at approximately 11 m depth. No temporal series of abiotic parameters were available for this
120 exact location. However, variations in the physico-chemical parameters (surface
121 measurements) at a station (called Estacade), located approximately 10 km from the Barre des
122 Flots site, were obtained from the *Service d’Observation des Milieux Littoraux* (SOMLIT)
123 between 2010 and 2013, with a sampling step of 15 days. Between October 2010 and March

124 2013, temperature varied between 8.1°C (January 2011) and 16.5°C (August 2011) with mean
125 values (\pm SE) of $10.1 \pm 0.2^\circ\text{C}$ in winter, $12.7 \pm 0.4^\circ\text{C}$ in spring and $15.8 \pm 0.02^\circ\text{C}$ in summer.

126 In Morlaix Bay (2009 to 2011), phytoplankton groups ($> 5\mu\text{m}$), the most important food
127 resource of *C. fornicata* (Decottignies et al. 2007), were mainly dominated by planktonic
128 diatoms in concentrations varying between 10 to 300 cells mL^{-1} (depending on the season) and
129 dinoflagellate species that were found at lower abundances (ca. 25 cells mL^{-1} ; Leroy 2011).

130

131 *Biological material*

132 *C. fornicata* forms stacks of several individuals in which each individual adheres to the
133 dorsal surface of the shell of the subjacent partner in the stack. It is a protandrous
134 hermaphrodite, meaning that the small individuals at the top of the stacks are generally males
135 and the large ones at the bottom, females (Coe 1936). After sampling, stacks were brought
136 directly to the *Station Biologique de Roscoff* where they were kept in natural, unfiltered
137 seawater for 6 weeks at a temperature gradually lowered to 10°C , reflecting the seasonal drop
138 in temperature between autumn and winter. Sexually mature individuals (more than 1 cm in
139 length) were selected and separated into two class sizes: small individuals (29.5 ± 0.9 mm
140 length) from the top of the stack and larger ones (45.4 ± 0.6 mm length) from the bottom. They
141 were separated from the stack and individually labeled with tags glued on their shell. Empty
142 subjacent shells, whose soft tissue was removed, served as substratum for the sampled live
143 individuals. Other empty shells whose size was similar to that of the substratum shell of live
144 individuals were also selected for flux corrections (see part “Metabolic rates and O:N ratios”
145 below). All the shells were gently brushed to remove epibionts without altering periostracum
146 layer.

147 Length (in mm), volume (in mL) and tissue dry weight (DW in g) of the live individuals
148 were determined for each incubated specimen at the end of the whole experiment. Length was

149 measured with calipers, volume was estimated as the volume of seawater moved when
150 individual was immersed and DW was determined after drying fresh samples at 60°C for 48 h.

151

152 *Experimental conditions*

153 Single small and large individuals, along with their substratum shell, were randomly
154 distributed into nine 10 L aquaria with 10 individuals of each class size per each aquarium.
155 Empty shells were also distributed into nine other 10 L aquaria (4 shells per aquarium). At the
156 beginning of the experiment, pH was gradually decreased over 2 weeks by 0.02 pH unit per day
157 from 8.1 until the different pH treatments were reached. *C. fornicata* individuals and empty
158 shells were then subsequently held for 24 weeks (12 January to 27 June 2012) in three $p\text{CO}_2$
159 treatments selected according to the recommendations in Barry et al. (2010): (1) 390 μatm (pH_T
160 = 8.07) represented current $p\text{CO}_2$, (2) 750 μatm (pH_T = 7.82) corresponded to the elevated $p\text{CO}_2$
161 level predicted by the IPCC for the end of the century (Solomon et al. 2007) and (3) 1400 μatm
162 (pH_T = 7.56) represented a $p\text{CO}_2$ five-fold higher than preindustrial $p\text{CO}_2$ (280 μatm) also
163 predicted for 2100 (Stocker et al. 2013). $p\text{CO}_2$ was adjusted by bubbling CO_2 -free air (current
164 $p\text{CO}_2$) or pure CO_2 (elevated $p\text{CO}_2$) in three 100 L header tanks supplied with unfiltered
165 seawater pumped directly from the foot of the *Station Biologique de Roscoff*. Each of the three
166 $p\text{CO}_2$ treatments had six replicate 10 L aquaria, three for live organisms and three for empty
167 shells. They continuously received CO_2 -treated seawater at a rate of 9 L h⁻¹ (i.e. a renewal rate
168 of 90% h⁻¹) from the header tanks. $p\text{CO}_2$ was monitored and controlled by an offline feedback
169 system (IKS Aquastar, Karlsbad, Germany) that regulated the addition of gas in the header
170 tanks. The pH values of the IKS system were adjusted from daily measurements of pH_T in the
171 18 aquaria using a pH meter (826 pH mobile, Metrohm AG, Herisau, Switzerland) calibrated
172 with Tris HCl and 2-aminopyridine HCl buffers (Dickson et al. 2007).

173 In each $p\text{CO}_2$ treatment, temperature was raised from 10 to 19°C with an incremental
174 step of 3°C. The first three temperature levels (10 to 16°C) simulated the natural change in
175 temperature from winter to summer in Morlaix Bay whereas the last level (19°C) corresponded
176 to a temperature increase of 3°C predicted for the end of the century (Solomon et al. 2007). *C.*
177 *fornicata* individuals were held for three weeks at each temperature before carrying out the
178 metabolic measurements (see below). This acclimation time was long enough to overcome the
179 immediate stress response (Meistertzheim et al. 2007). Temperature was maintained at (1) 10°C
180 (1st trial period) from 16 January to 12 February 2012; (2) 13°C (2nd trial period) from 27
181 February to 25 March 2012; (3) 16°C (3rd trial period) from 9 April to 6 May 2012, and (4)
182 19°C (4th trial period) from 21 May to 27 June 2012. Between two temperature levels,
183 temperature was gradually increased by 0.2°C day⁻¹ over two weeks. The 18 aquaria were
184 placed in thermostatic baths in which temperature was regulated to within $\pm 0.2^\circ\text{C}$ using
185 submersible 150 to 250 W heaters controlled by the IKS system.

186 Three independent 10 L aquaria named “control” were maintained at 10°C under
187 ambient pH (with no $p\text{CO}_2$ control) until the end of the experiment in order to estimate a
188 potential bias on metabolism induced by the mesocosm experiment over time. Each aquarium
189 contained 10 small and 10 large slipper limpets on their substratum shell and was supplied with
190 the same seawater sourced from the header tanks. They were kept in a thermostatic bath
191 regulated at 10°C by an aquarium chiller (TC5, TECO®, Ravenna, Italy).

192 In addition to the natural phytoplankton found in the unfiltered seawater, all slipper
193 limpets were fed twice a week with a stock solution composed of the diatom *Chaetoceros*
194 *gracilis* ($\sim 15 \times 10^6$ cells mL⁻¹) and the dinoflagellate *Isochrysis affinis galbana* ($\sim 26 \times 10^6$
195 cells mL⁻¹); 400 mL of this microalgal mix was added to each aquarium at each feeding.
196 Seawater flow was stopped for 2 h when organisms were fed and filtering actively. During this
197 feeding time, pH variation did not exceed 0.05 units.

198 Individuals that did not adhere to their substratum shell and that showed no reaction
199 when their foot was stimulated were counted as dead and removed from the tanks. Mortality
200 reached only 8% at the end of the experiment among all $p\text{CO}_2$ conditions.

201

202 *Seawater parameter monitoring*

203 Seawater parameters were monitored throughout the experiment. pH_T and temperature
204 were recorded daily in each of the 21 aquaria (18 + 3 controls) using a pH meter (826 pH
205 mobile, Metrohm AG, Herisau, Switzerland) as described above. Total alkalinity (A_T) was
206 measured at each trial period by 0.01 N HCl potentiometric titration on an automatic titrator
207 (Titroline alpha, Schott SI Analytics, Mainz, Germany). Salinity was also measured at each trial
208 period with a conductimeter (LF 330/ SET, WTW, Weilheim, Germany). Seawater carbonate
209 chemistry, i.e. dissolved inorganic carbon (DIC), $p\text{CO}_2$ and the saturation state of aragonite
210 (Ω_{Ar}) were calculated for each $p\text{CO}_2$ level and temperature with CO₂SYS software (Lewis and
211 Wallace 1998) using constants from Mehrbach et al. (1973) refitted by Dickson & Millero
212 (1987).

213

214 *Metabolic rates and O:N ratios*

215 Metabolic rates were assessed at each temperature level after a four-day starvation
216 period and after the shells were gently cleaned to remove biofilm-forming organisms. Two
217 small and two large individuals were selected per aquarium. They were incubated individually
218 in 185 mL (small) and 316 mL (large) acrylic chambers (Engineering & Design Plastics Ltd,
219 Cambridge, UK) filled with seawater from their respective aquaria. They were put on a plastic
220 grid above a stirring bar (speed 100 rpm.), which ensured water homogeneity. Chambers were
221 placed in their original aquaria for incubation to keep the temperature constant. Incubations
222 were carried out in dark for 2 to 10 h, depending on temperature and limpet size, to maintain

223 oxygen saturation above 80% until the end of the incubation. At each temperature period, empty
224 shell incubations were carried out to correct individual rates for fluxes related to the substratum
225 shell. Blank incubations containing only seawater from the aquarium also helped to correct
226 fluxes for any microbiological activity in seawater.

227 Oxygen concentrations were measured at the beginning and the end of the incubation
228 period with a non-invasive fiber-optics system and reactive oxygen spots attached to the inner
229 wall of the chambers (FIBOX 3, PreSens, Regensburg, Germany). Spots were calibrated at the
230 beginning of each trial period with 0% and 100% oxygen buffers. Seawater was sampled for
231 ammonium (NH_4^+) concentration and A_T measurements with 100 mL syringes at the beginning
232 of the incubation, directly in the aquaria just after the chambers were closed, and at the end of
233 the incubation, in the incubation chamber itself. Samples were filtered through 0.7 μm
234 Whatman GF/F filters into 100 mL glass bottles and fixed with reagent solutions for ammonium
235 or poisoned with mercuric chloride (0.02% vol/vol; Dickson et al. 2007) for A_T measurements.
236 Vials were stored in the dark pending analysis. NH_4^+ concentrations were then determined using
237 the Solorzano method (Solorzano 1969) based on spectrophotometry at a wavelength of 630
238 nm (spectrophotometer UV-1201V, Shimadzu Corp, Kyoto, Japan). A_T (in $\mu\text{Eq L}^{-1}$) values
239 were determined by 0.01 N HCl potentiometric titration on an automatic titrator (Titroline
240 alpha, Schott SI Analytics, Mainz, Germany) and by using the Gran method (non-linear least-
241 squares fit) applied to pH values from 3.5 to 3.0 (Dickson et al. 2007).

242 Respiration (in $\mu\text{mol O}_2 \text{ g}^{-1} \text{ DW h}^{-1}$; equation [1]) and excretion (in $\mu\text{mol NH}_4^+ \text{ g}^{-1} \text{ DW}$
243 h^{-1} ; equation [2]) were directly calculated from oxygen and ammonium concentrations,
244 respectively. Net calcification (in $\mu\text{mol CaCO}_2 \text{ g}^{-1} \text{ DW h}^{-1}$; equation [3]) was estimated using
245 the alkalinity anomaly technique (Smith and Key 1975) based on a decrease in A_T by 2
246 equivalents for each mole of CaCO_3 precipitated (Wolf-Gladrow et al. 2007). As ammonium

247 production increases alkalinity in a mole-per-mole ratio (Wolf-Gladrow et al. 2007), the
248 alkalinity variation was corrected by the ammonium flux to calculate CaCO₃ fluxes.

249 [1] $R = \frac{\Delta O_2 \times V}{\Delta t \times DW}$

250 [2] $E = \frac{\Delta NH_4^+ \times V}{\Delta t \times DW}$

251 [3] $G_n = -\frac{(\Delta A_T - \Delta NH_4^+) \times V}{2 \times \Delta t \times DW}$

252 where ΔO_2 (in $\mu\text{mol O}_2 \text{ L}^{-1}$) is the difference between initial and final O₂ concentrations; Δ
253 NH₄⁺ (in $\mu\text{mol NH}_4^+ \text{ L}^{-1}$) is the difference between initial and final NH₄⁺ concentrations; ΔA_T
254 is the difference between initial and final total alkalinity ($\mu\text{mol Eq L}^{-1}$); V (in L) is the volume
255 of the chamber minus *C. fornicata* volume; Δt (in h) is the incubation time and DW (in g) is the
256 soft tissue dry weight of incubated *C. fornicata*.

257 In addition, oxygen consumption of the individuals maintained at 10°C during the
258 experiment were assessed on six small and six large individuals at each trial period, following
259 the technique described above. These “controls” tested if mesocosm conditioning induced
260 metabolic stress over time.

261 The O:N ratio, which corresponds to the amount of oxygen consumed for nitrogen
262 excreted, was calculated from respiration and excretion rates except for the experiments run at
263 10°C for which rates were too low to obtain significant data. Generally, the O:N ratio is
264 considered a common indicator of the proportion of the three metabolic substrates
265 (carbohydrates, lipids and proteins) used in energy metabolism (Mayzaud and Conover 1988).
266 The atomic ratio of oxygen uptake and excreted nitrogen was calculated following the equation
267 [4] based on Thomsen & Melzner (2010):

268 [4]: $O:N = R / E$

269 where R is the respiration rate used as a proxy of the quantity of oxygen consumed by the
270 individual and E, the excretion rate representing the concentration of nitrogen excreted.

271

272 *Filtration rates*

273 At each trial period, the filtration rate of three small and three large slipper limpets per
274 $p\text{CO}_2$ condition (i.e. 1 individual per size per aquarium) was determined by calculating
275 clearance rates (Coughlan 1969). To do so, 10 and 20 mL of a microalgae mix (*C. gracilis*, *T.*
276 *affinis galbana*, 1:1) were added to the small and large chambers (same as for metabolic
277 measurements), respectively, using a 10 mL syringe equipped with a thin tube. The mean initial
278 concentration of the mix was $1\,200\,000 \pm 310\,000$ cell mL^{-1} . In parallel, control incubations
279 containing only microalgae were carried out to check that phytoplankton cells did not multiply
280 significantly during the incubation. Water from the chambers was sampled with the syringe
281 every 15 min until the water became totally clear (around 2 h). Samples were immediately fixed
282 with 25% glutaraldehyde and frozen at -80°C pending analyses (Marie et al. 1999). The number
283 of microalgal cells in each sample was then determined on 200 μL aliquots using flux cytometry
284 (Cell Lab QuantaTM, SC, Beckman Coulter, USA). Filtration rates (F, in $\text{mL SW g}^{-1} \text{DW min}^{-1}$)
285 were calculated following equation [5]:

$$286 \quad [5] \quad F = V \times \frac{\ln[C_i] - \ln[C_f]}{\Delta t \times \text{DW}}$$

287 where $[C_i]$ and $[C_f]$ (in cell mL^{-1}) were respectively the initial and final cell concentrations in
288 the chamber water; V (in L) is the volume of the chamber minus individual *C. fornicata* volume;
289 Δt (in h) is the incubation time and DW (in g) is the tissue dry weight of the individual
290 incubated.

291

292 *Statistical analyses*

293 All statistical analyses were performed using the R software, version 2.15.0 (R Core
294 Team 2013). Normality and homoscedasticity were checked using Kolmogorov-Smirnov's test
295 and Levene's test, respectively, before each statistical test. Spatial pseudoreplication effect was

296 first tested by considering “aquarium” as a random factor (p -value < 0.05). Then, statistical
297 analyses were simplified to two-way ANOVAs with repeated measurements on the same
298 individual through the four trial periods (different temperature levels) separately for small and
299 large individuals. These analyses were performed for the four physiological rates (respiration,
300 excretion, calcification and filtration) and the O:N ratio, assuming $p\text{CO}_2$ and temperature as
301 fixed factors. Student-Newman-Keuls (SNK) post hoc tests were applied to identify differences
302 among treatments with a confidence level of 95% when ANOVA showed significant results. In
303 parallel, any changes in the respiration rate of individuals constantly maintained at 10°C
304 through time were assessed using a non-parametric Friedman test for repeated measurements,
305 separately for small and large slipper limpets. All results are given as mean \pm standard error
306 (SE).

307

308 **Results**

309

310 *Seawater parameters*

311 The mean temperature and carbonate chemistry parameters among the $p\text{CO}_2$ and
312 temperature conditions are presented in Table 1. Temperature was stable at each trial period
313 with a variability lower than 0.5°C . The different $p\text{CO}_2$ levels remained close to the selected
314 values of 390, 750 and $1400\ \mu\text{atm}$ except at 19°C where all $p\text{CO}_2$ increased from the baseline
315 ($+ 100$ - $200\ \mu\text{atm}$). A_T ranged from 2365 ± 2 to $2422 \pm 2\ \mu\text{Eq kg}^{-1}$. Ω_{Ar} decreased by less than
316 1 only in the $1400\ \mu\text{atm } p\text{CO}_2$ condition. Salinity varied between 34.2 ± 0.1 and 35.1 ± 0.1
317 among the different $p\text{CO}_2$ and temperature levels with no effect of the temperature increase on
318 salinity.

319

320 *Respiration, excretion and O:N ratio*

321 Respiration and excretion rates changed significantly with temperature, but not with
322 $p\text{CO}_2$, in small and large individuals (Figure 1, Table 2). After pooling results for all $p\text{CO}_2$
323 conditions, mean respiration rates in small *C. fornicata* increased from 3.78 $\mu\text{mol O}_2 \text{g}^{-1} \text{DW h}^{-1}$
324 at 10°C to 11.76 $\mu\text{mol O}_2 \text{g}^{-1} \text{DW h}^{-1}$ at 19°C. In large individuals, the lowest mean respiration
325 rate was recorded at 10°C (4.82 $\mu\text{mol O}_2 \text{g}^{-1} \text{DW h}^{-1}$) whereas rates did not differ from 13 to
326 19°C with a mean value of 11.50 $\mu\text{mol O}_2 \text{g}^{-1} \text{DW h}^{-1}$. Oxygen fluxes measured on empty shells
327 represented only 4% of the whole organism fluxes measured and decreased only slightly with
328 temperature.

329 Mean excretion rates calculated among $p\text{CO}_2$ conditions for small *C. fornicata*
330 individuals gradually increased from 0.15 $\mu\text{mol NH}_3 \text{g}^{-1} \text{DW h}^{-1}$ at 10°C to 1.47 $\mu\text{mol NH}_3 \text{g}^{-1}$
331 DW h^{-1} at 19°C. Excretion rates of large individuals showed a parabolic trend with an increase
332 from 10°C (0.16 $\mu\text{mol NH}_3 \text{g}^{-1} \text{DW h}^{-1}$) to 16°C (1.34 $\mu\text{mol NH}_3 \text{g}^{-1} \text{DW h}^{-1}$) followed by a
333 decrease at 19°C (0.74 $\mu\text{mol NH}_3 \text{g}^{-1} \text{DW h}^{-1}$). The ammonium fluxes of empty shells
334 represented less than 1% of the fluxes estimated for whole organisms and were higher at 10°C
335 than at the other temperature levels (rates practically nil).

336 O:N ratios varied greatly, ranging from 2.86 to 31.68 with a mean value of 12.91 ± 0.56 .
337 They varied with $p\text{CO}_2$ or temperature according to size (Table 2, Figure 2). In small *C.*
338 *fornicata* individuals, O:N ratios were the highest at 750 μatm and similar between 380 and
339 1400 μatm . In large individuals, the O:N ratios varied with temperature and were significantly
340 higher at 16°C.

341

342 *Filtration*

343 Temperature significantly affected filtration rates in both small and large individuals
344 (Figure 1, Table 2). In small *C. fornicata*, mean filtration rates among $p\text{CO}_2$ were similar
345 between 10 and 16°C (25.50 $\text{mL g}^{-1} \text{DW min}^{-1}$), but increased at 19°C (54.30 $\text{mL g}^{-1} \text{DW min}^{-1}$)

346 ¹). $p\text{CO}_2$ alone did not affect the filtration rate but the interaction of $p\text{CO}_2$ and temperature was
347 significant (Table 2, p-value < 0.001). At 19°C, filtration rates increased significantly with the
348 increase in $p\text{CO}_2$. In large individuals, mean filtration rates increased gradually from 10°C (5.43
349 $\text{mL g}^{-1} \text{DW min}^{-1}$) to 19°C (25.78 $\text{mL g}^{-1} \text{DW min}^{-1}$) without any effect of $p\text{CO}_2$ conditions.

350

351 *Calcification*

352 Calcification rates were significantly affected by $p\text{CO}_2$ increase in both small and large
353 individuals but not by temperature (Figure 1, Table 2). Pooling all temperature levels together,
354 mean calcification rates were similar at $p\text{CO}_2$ of 390 μatm (1.88 and 1.63 $\mu\text{mol CaCO}_3 \text{ g}^{-1} \text{DW}$
355 h^{-1} in small and large individuals, respectively) and 750 μatm (1.02 and 0.60 $\mu\text{mol CaCO}_3 \text{ g}^{-1}$
356 DW h^{-1} in small and large, respectively), but significantly lower at 1400 $\mu\text{atm } p\text{CO}_2$ (-2.53 and
357 -1.77 $\mu\text{mol CaCO}_3 \text{ g}^{-1} \text{DW h}^{-1}$ in small and large individuals, respectively). In the highest $p\text{CO}_2$
358 condition (1400 μatm), net calcification rates were negative, corresponding to dissolution.
359 Although the interaction between $p\text{CO}_2$ and temperature was not significant for either small or
360 large limpets, $p\text{CO}_2$ response appeared to vary as a function of temperature, particularly at 1400
361 μatm . In this drastic $p\text{CO}_2$ condition, organisms globally dissolved at 10, 13 and 16°C and
362 calcified (or dissolved less) at 19°C.

363 Calcification rates decreased with the decrease in the mean aragonite saturation state
364 (Ω_{Ar}) which correlated with $p\text{CO}_2$ increase (Figure 3). When Ω_{Ar} decreased below the threshold
365 of 1, calcification rates were always negative reflecting a dissolution process. At the 750 and
366 1400 $\mu\text{atm } p\text{CO}_2$ conditions, Ω_{Ar} was higher at 19°C than at the other temperature levels
367 because the saturation state increases with temperature.

368

369 *Mesocosm controls*

370 In the aquaria maintained at 10°C throughout the entire experiment, temperature was
371 stable over the first weeks of the experiment and slowly increased from 8 April to the end of
372 the experiment until reaching a mean of 12.4°C between 21 April and 15 June because we had
373 technical problems with the chiller (Table 1). Respiration in small individuals showed high
374 variation over time (Figure 4, white bars) but no time effect was detected (Friedman test, $df =$
375 3 , $\chi^2 = 6.6$, $p = 0.086$, $n = 6$). Conversely, respiration rates of large individuals increased
376 throughout the experiment (Figure 4, gray bars) with a significant time effect (Friedman test,
377 $df = 3$, $\chi^2 = 9.4$, $p = 0.024$, $n = 6$).

378

379 **Discussion**

380 An increase in temperature affected three of the four physiological processes assessed
381 on small and large *C. fornicata* individuals. In particular, respiration and ammonia excretion
382 rates clearly increased along the tested temperature gradient. In contrast, increases in $p\text{CO}_2$
383 affected only net calcification of the slipper limpets. Interestingly, the coupled effect of
384 temperature and $p\text{CO}_2$ improved the rate of calcification in the most drastic conditions,
385 particularly in small individuals.

386

387 *Temperature effect*

388 The respiration ($0.6 - 34.6 \mu\text{mol O}_2 \text{ g}^{-1} \text{ DW h}^{-1}$) and excretion rates ($-2 - 4.4 \mu\text{mol NH}_3$
389 $\text{g}^{-1} \text{ DW h}^{-1}$) measured at $390 \mu\text{atm } p\text{CO}_2$ in small and large *C. fornicata* individuals ranged
390 metabolic rates recorded *in situ* in the Bay of Brest in northwestern France (4 to $45 \mu\text{mol O}_2 \text{ g}^{-1}$
391 DW h^{-1} and 0.5 to $2.3 \mu\text{mol NH}_3 \text{ g}^{-1} \text{ DW h}^{-1}$; Martin et al. 2006). Both rates increased with
392 temperature in small and large individuals regardless of $p\text{CO}_2$. Although respiration rates
393 gradually increased with temperature in small *C. fornicata* individuals, they only increased from
394 10°C to 13°C , remaining stable at higher temperatures in large *C. fornicata*. This increase is a

395 common response due to the rate-enhancing effects of temperature on physiological and
396 biochemical reactions in ectotherms (Cossins and Bowler 1987). The intensity of respiratory
397 and excretory processes were also dependent of body size. The respiration and excretion rates
398 of small individuals were higher than those of large individuals because the metabolic rate (per
399 unit biomass) decreases with increasing individual size (Parsons et al. 1984; Von Bertalanffy
400 1951). Small individuals have higher energy consumption because they grow faster than the
401 large individuals (Von Bertalanffy 1964).

402 The filtration rates measured in small and large *C. fornicata* fall into the range of
403 maximum feeding rates calculated by Newell and Kofoed (1977) in *C. fornicata* between 11
404 and 20°C (18 to 41 mL g⁻¹ min⁻¹; 15°C acclimated individuals). Rates were higher in small than
405 in large individuals because, again, small organisms feed more actively per unit body mass
406 (Sylvester et al. 2005). Filtration rates increased with temperature as previously described in
407 other studies (Newell and Kofoed 1977). In small individuals, rates were constant between 10
408 and 16°C and increased only at 19°C while they increased regularly with temperature in the
409 large individuals. In Calyptraeidae, small individuals — i.e. males with low mobility — utilize
410 two feeding strategies: grazing with radula and filtration with gills (Navarro and Chaparro
411 2002). Therefore, small individuals may have supplemented their diet between 10 and 16°C by
412 grazing. For the increased energy requirements at 19°C, small slipper limpets may also increase
413 their filtration rate to meet these supplementary needs. In large sedentary individuals (usually
414 females), filtration is the only feeding mechanism (Navarro and Chaparro 2002) and filtration
415 rate increases with temperature to help cover the higher energy needs.

416 Surprisingly, temperature did not affect calcification rates although an increase was
417 expected in response to the increase in metabolism and energy requirements (Martin et al.
418 2006). Because mollusk shell production is an energetically costly process (Gazeau et al. 2013),
419 the absence of any change in calcification rates may be due to food limitation during the

420 experiment, especially at elevated temperatures (16 and 19°C). At these temperatures,
421 providing additional food only twice a week may not have been sufficient to support maximal
422 individual shell growth under pH stressful conditions. If food had been provided more regularly
423 and/or in higher quantities, *C. fornicata* calcification may not be potentially restricted and
424 individuals may have better mitigated the effect of high $p\text{CO}_2$ (Thomsen et al. 2014). Future
425 experiments should include measuring integrated shell growth at each temperature level to
426 determine the food effect more completely.

427 Mesocosm experiments cannot perfectly reproduce *in situ* conditions such as natural
428 diet or tidal cycles. This may lead to an increased stress for the organisms grown in these
429 systems (Bibby et al. 2008). The mesocosm effect on organisms was tested through O_2
430 consumption measurements in individuals kept at constant temperature throughout the
431 experiment (“controls”). The respiration rates did not change over time in small individuals,
432 whereas the respiration in large individuals increased slightly in correlation with a +2°C
433 temperature increase from the beginning to the end of the experiment, because of technical
434 problems with the chiller. Although food may have constituted a bias, particularly in the one-
435 off calcification response to temperature, the absence of strong changes in respiration rates in
436 “controls”, unexceptional metabolic rates ranging those measured *in situ* and very low mortality
437 during the experiment (only 8%) all suggest the absence of any acute mesocosm effect on the
438 other physiological traits of *C. fornicata*.

439

440 *pCO₂ effect*

441 In contrast to temperature, $p\text{CO}_2$ did not affect *C. fornicata* respiration or excretion rates
442 regardless of size. Other studies have underlined a lack of any $p\text{CO}_2$ effect on bivalve and limpet
443 respiration (Dickinson et al. 2012; Fernandez-Reiriz et al. 2012; Marchant et al. 2010), although
444 some mollusk species exposed to high $p\text{CO}_2$ levels have shown metabolic depression (i.e.

445 decrease in oxygen uptake) to compensate — albeit often drastic — $p\text{CO}_2$ increases
446 (Michaelidis et al. 2005; Navarro et al. 2013). Responses of ammonia excretion to high $p\text{CO}_2$
447 in mollusks are also specific: increase in ammonia excretion can occur under elevated $p\text{CO}_2$
448 (Fernandez-Reiriz et al. 2011; Langenbuch and Pörtner 2002; Range et al. 2014) while some
449 bivalves show opposing trends (Liu and He 2012; Navarro et al. 2013). The increase in
450 ammonia excretion under increased $p\text{CO}_2$ conditions can be interpreted as an internal pH
451 regulatory mechanism, sometimes based on protein catabolism (Fernandez-Reiriz et al. 2012;
452 Thomsen and Melzner 2010). In our study, neither change in excretion rates nor in O:N ratios
453 calculated were detected between the 390 and 1400 μatm conditions. This similarity indicates
454 that potential intracellular pH regulation of *C. fornicata* was not induced by enhancing protein
455 metabolism (Fernandez-Reiriz et al. 2012). Thus, the potential for metabolic resistance of *C.*
456 *fornicata* to elevated $p\text{CO}_2$ is likely due to another effective acidosis-buffering system, such as
457 the increase in internal HCO_3^- concentrations (Gutowska et al. 2010; Michaelidis et al. 2005)
458 or higher H^+ excretion (Melzner et al. 2009; Pörtner et al. 2005).

459 Similarly to the respiration and excretion processes, filtration rates did not change as a
460 function of $p\text{CO}_2$ in either small or large *C. fornicata* in our study. Filtration responses with
461 respect to $p\text{CO}_2$ depend most of the time on the presence of metabolic depression (Fernandez-
462 Reiriz et al. 2011; Liu and He 2012; Navarro et al. 2013). The absence of variation in filtration
463 rates at the different $p\text{CO}_2$ levels indicates that the quantity of food ingested by *C. fornicata* did
464 not vary either. Food is known to interact with other stressors, such as $p\text{CO}_2$, and significantly
465 influence metabolic responses (Melzner et al. 2011; Pansch et al. 2014). Quality or quantity
466 changes in the diet can even worsen the condition of invertebrates (Berge et al. 2006; Vargas et
467 al. 2013). Although our microalgal mix did not perfectly match the natural diet of *C. fornicata*
468 (Barillé et al. 2006; Decottignies et al. 2007), the diatoms and dinoflagellate microalgae
469 provided in the experiment correspond to the main taxa present in Morlaix Bay, assuming a

470 nutritional quality close to the natural diet. However, we cannot assure that the quantity of food
471 was not a limiting factor in our experiment. To be sure that microalgae supplied would not
472 represent a bias, the slipper limpets should be fed *ad libitum* which represented a technical issue
473 on a 6 month experiment.

474 In our study, net calcification was similar between 390 and 750 $\mu\text{atm } p\text{CO}_2$ and strongly
475 decreased at 1400 $\mu\text{atm } p\text{CO}_2$ regardless of size, which is a common response in mollusks
476 (Beniash et al. 2010; Melatunan et al. 2013; Range et al. 2011). This pattern contrasts with that
477 reported in Ries et al. (2009), with a parabolic response in *C. fornicata* calcification with the
478 highest rates observed at 600 $\mu\text{atm } p\text{CO}_2$. The stability of calcification rate at 750 $\mu\text{atm } p\text{CO}_2$
479 (compared to 390 $\mu\text{atm } p\text{CO}_2$) may be due to the biological control of the calcification process
480 and/or the presence of the periostracum, the organic layer covering the crystalline layers of the
481 shell. This organic layer has been predicted to play a great role in maintaining shell integrity of
482 mollusks in elevated $p\text{CO}_2$ conditions (Ries et al. 2009) and to protect them from dissolution in
483 CaCO_3 -undersaturated waters (Huning et al. 2013). Moreover, mollusks may be able to
484 maintain extrapallial fluid in chemical conditions favoring CaCO_3 precipitation at the
485 calcification site, even if external seawater $p\text{CO}_2$ is high (Hiebenthal et al. 2013). Regulation
486 of enzymes involved in the calcification process, such as chitinase (Cummings et al. 2011) or
487 carbonic anhydrase (Ivanina et al. 2013), may also help maintain calcification in high $p\text{CO}_2$
488 conditions. In our study, at 1400 μatm , calcification rates dropped, perhaps due to physiological
489 changes in the internal acid-base balance affecting shell deposition (Waldbusser et al. 2011) or
490 to an eroded and/or damaged periostracum (pers. obs.). Degradation of this protective layer may
491 lead to higher vulnerability of the shell to external dissolution processes (Range et al. 2012;
492 Ries et al. 2009), which occurs not only in dead shells but also in live animals (Harper 2000).
493 Furthermore, chemical dissolution increased with an increase in $p\text{CO}_2$ and a correlated decrease

494 in Ω_{Ar} ; the combined effect led to a decrease in net calcification rates observed in both small
495 and large *C. fornicata* individuals at high pCO_2 conditions.

496

497 *Combined effects of temperature and pCO_2*

498 In the range of pCO_2 and temperatures tested, the interaction of these two variables had
499 no negative effect on *C. fornicata* respiration and excretion rates. As a eurythermal species even
500 coping with high temperature in some bays during summer (e.g. Bassin d’Arcachon in
501 southwestern France; De Montaudouin et al. 1999), *C. fornicata* can have an optimal
502 temperature of 19°C or higher (Diederich and Pechenik 2013; Noisette et al. 2015). Thus, 19°C
503 may not constitute a real thermal stress and not transgress the metabolic optimal threshold for
504 this species. Increase in temperature is predicted to enhance sensitivity to high pCO_2 levels
505 beyond the optimal temperature of the species and close to its upper limit of thermal tolerance
506 (Pörtner and Farrell 2008). However, at the cold side of a species optimal temperature, warming
507 can increase resilience to ocean acidification (Gianguzza et al. 2014). Therefore, an increase in
508 temperature may actually improve tolerance to pCO_2 increases in *C. fornicata*.

509 Calcification rates of both small and large *C. fornicata* showed a positive trend with
510 temperature in the most drastic pCO_2 conditions (1400 μatm). Temperature-mediated increases
511 in metabolism and feeding rates may potentially offset reductions in calcification rates caused
512 by high pCO_2 conditions (Melzner et al. 2011; Thomsen et al. 2014). In addition to this
513 physiological effect, moderate warming can mediate the effects of ocean acidification by the
514 chemical effect on seawater chemistry (Kroeker et al. 2014). Temperature affects CO_2 solubility
515 in seawater as well as the equilibrium coefficients governing carbonate chemistry (Millero
516 2007). As shown in our study, the saturation state of aragonite was greater in warmer water
517 than in colder water for a given pCO_2 , thereby enhancing calcification and reducing the
518 dissolution processes in the high pCO_2 conditions. These results highlight the importance of

519 considering the physiological and geochemical interactions between temperature and carbonate
520 chemistry when interpreting species' vulnerability to ocean acidification. A better
521 understanding of how warming influences species' responses to high $p\text{CO}_2$ levels through both
522 direct (e.g. increases in metabolic rates) and indirect pathways (e.g. changes in carbonate
523 chemistry) is thus necessary.

524

525 *Conclusion*

526 A trade-off between stressors may affect the physiology of organisms in an unexpected
527 way (Kroeker et al. 2014). In our case, *C. fornicata* appeared to be able to tolerate slight
528 increases in $p\text{CO}_2$ but its calcification was affected by drastic conditions with a positive effect
529 of temperature, thereby mitigating any ocean acidification effects. This outcome highlights the
530 need of multistressor studies to understand the future of marine species in a context of climate
531 change in which different physico-chemical factors vary in different ways. Furthermore, our
532 results indicate that some species can be highly tolerant to future $p\text{CO}_2$ increases. *C. fornicata*
533 tolerance likely stems from mechanisms that allow it to acclimate or adapt to environmental
534 fluctuations in its habitat (Clark et al. 2013), because species living in environments with large
535 abiotic variations tend to have high phenotypic plasticity, allowing them to survive in stressful
536 conditions (Somero 2010). This capacity to resist decreases in pH may reinforce the ecological
537 role of *C. fornicata* populations in the ecosystems in which they are established, even under
538 projected future conditions anticipated due to climate change.

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Tables

Table 1: Mean seawater temperature and parameters of the carbonate system in each $p\text{CO}_2$ treatment (3 aquaria per treatment) and at each trial period (i.e. temperature level). The pH_T (pH on the total scale) and total alkalinity (A_T) were measured whereas the other parameters were calculated. Mean A_T calculated for each trial period ($n = 3$ for controls 10°C and $19 < n < 30$ for other condition $p\text{CO}_2$ conditions) and $p\text{CO}_2$ condition was used for the calculations. $p\text{CO}_2$, CO_2 partial pressure; DIC, dissolved inorganic carbon and Ω_{Ar} , saturation state of seawater with respect to aragonite.

	n	Temperature ($^\circ\text{C}$)		pH_T		$p\text{CO}_2$ (μatm)		A_T ($\mu\text{Eq kg}^{-1}$)		DIC ($\mu\text{mol C kg}^{-1}$)		Ω_{Ar}	
		Mean	SE	Mean	SE	Mean	SE	Mean	SE	Mean	SE	Mean	SE
1st trial period (10°C)													
390 μatm	23	9.7	0.2	8.14	0.01	322	7	2365	2	2138	4	2.47	0.04
750 μatm	23	9.8	0.2	7.82	0.01	729	19	2368	2	2270	4	1.33	0.03
1400 μatm	23	9.5	0.2	7.55	0.03	1486	75	2376	2	2366	11	0.78	0.08
control 10°C	40	9.2	0.2	8.19	0.02	288	17	2370	3	2115	8	2.73	0.07
2nd trial period (13°C)													
390 μatm	27	12.9	0.2	8.12	0.02	356	25	2418	2	2167	8	2.76	0.07
750 μatm	27	13.0	0.1	7.81	0.01	781	20	2416	2	2304	3	1.48	0.03
1400 μatm	27	12.8	0.1	7.53	0.01	1557	43	2422	2	2405	4	0.82	0.02
control 10°C	41	11.0	0.1	8.18	0.01	297	12	2419	2	2152	5	2.88	0.05
3rd trial period (16°C)													
390 μatm	28	15.9	0.1	8.08	0.01	376	10	2379	5	2126	5	2.80	0.05
750 μatm	28	16.1	0.1	7.82	0.00	748	8	2369	5	2238	2	1.66	0.01

1400 μatm	28	16.0	0.1	7.55	0.01	1492	19	2380	5	2345	2	0.94	0.01
control 10°C	42	11.4	0.1	8.23	0.01	253	6	2376	4	2083	5	3.13	0.05

4th trial period (19°C)

390 μatm	23	18.4	0.5	8.02	0.01	450	10	2391	2	2152	5	2.70	0.05
750 μatm	23	18.6	0.5	7.77	0.01	858	19	2395	3	2266	4	1.68	0.04
1400 μatm	23	18.4	0.5	7.51	0.01	1652	41	2394	2	2359	4	0.96	0.03
control 10°C	23	12.4	0.1	8.20	0.01	280	12	2393	1	2107	8	3.07	0.08

1 **Table 2:** Summary of two-way repeated measurements ANOVAs followed by Student-Newman-Keuls post hoc tests testing the effect of $p\text{CO}_2$,
 2 temperature and their interaction on *Crepidula fornicata* physiology. Numbers in bold indicate significant p-values and values with different letters
 3 are significantly different at $p < 0.05$.

	Two-way repeated measurements ANOVAs									Post hoc SNK tests						
	Factors			Factors			Factors			Factors						
	$p\text{CO}_2$		Temperature		$p\text{CO}_2 \times \text{Temperature}$		$p\text{CO}_2$ (μatm)			Temperature ($^\circ\text{C}$)						
df	F	p	df	F	p	df	F	p	390	750	1400	10	13	16	19	
Small individuals																
Respiration	2	1.685	0.219	3	14.530	< 0.001	6	1.893	0.103				a	b	b	c
Excretion	2	0.386	0.686	3	5.840	0.002	6	1.257	0.296				a	a,b	b	b
Filtration	2	0.271	0.766	3	15.439	< 0.001	6	5.996	< 0.001				a	a	a	b
Net calcification	2	6.705	0.008	3	1.849	0.152	6	2.307	0.050	a	a	b				
O:N ratio	2	4.944	0.022	2	2.214	0.127	4	0.382	0.819	a	b	a				
Large individuals																
Respiration	2	0.377	0.692	3	8.398	< 0.001	6	0.523	0.788				a	b	b	b
Excretion	2	0.563	0.581	3	17.850	< 0.001	6	0.371	0.893				a	b	c	b
Filtration	2	1.593	0.236	3	19.311	< 0.001	0	2.012	0.083				a	b	b	c
Net calcification	2	13.615	< 0.001	3	0.878	0.459	6	0.911	0.496	a	a	b				
O:N ratio	2	0.739	0.494	2	20.714	< 0.001	4	1.728	0.170				-	a	b	a

4

5 **Figures**

6

7 **Figure 1:** Individual respiration, ammonia excretion, filtration and net calcification rates in the three $p\text{CO}_2$
8 treatments (shaded in grey) at 10, 13, 16 and 19°C for small (< 3 cm in length) and large (> 4.5 cm in
9 length) *C. fornicata* individuals. Different letters above bars or before $p\text{CO}_2$ caption indicate significant
10 differences between temperature or $p\text{CO}_2$ conditions, respectively. Results are expressed as mean \pm
11 standard error, n = 6 individuals.

12

13 **Figure 2:** O:N ratios for the three $p\text{CO}_2$ treatments (shaded in grey) at 13, 16 and 19°C for small and large
14 *C. fornicata* individuals. Different letters above bars or before $p\text{CO}_2$ caption indicate significant differences
15 between temperature or $p\text{CO}_2$ conditions, respectively. Results are expressed as mean \pm standard error, n =
16 6 individuals.

17

18 **Figure 3:** Mean net calcification rates as function of aragonite saturation state, in the three $p\text{CO}_2$ treatments
19 (shaded in grey), at 10 (○), 13 (△), 16 (□) and 19°C (◇) for all *C. fornicata* individuals (n = 12
20 individuals).

21

22 **Figure 4:** Respiration rates in the control treatment (10°C) for the different trial periods (i.e. temperature
23 levels) for single small (white bars) and large (grey bars) *C. fornicata* individuals. Results are expressed as
24 mean \pm standard error, n = 6 individuals.

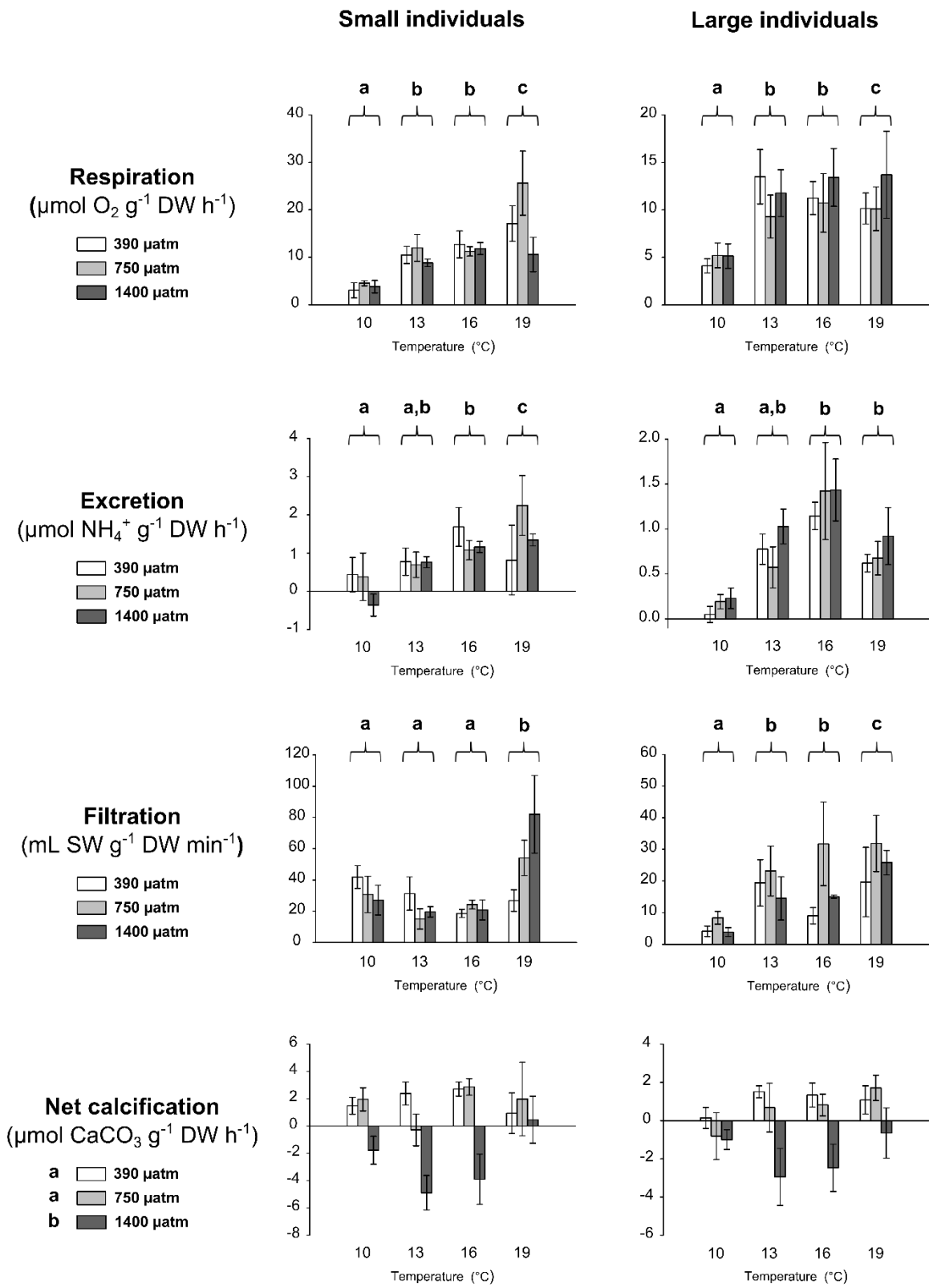
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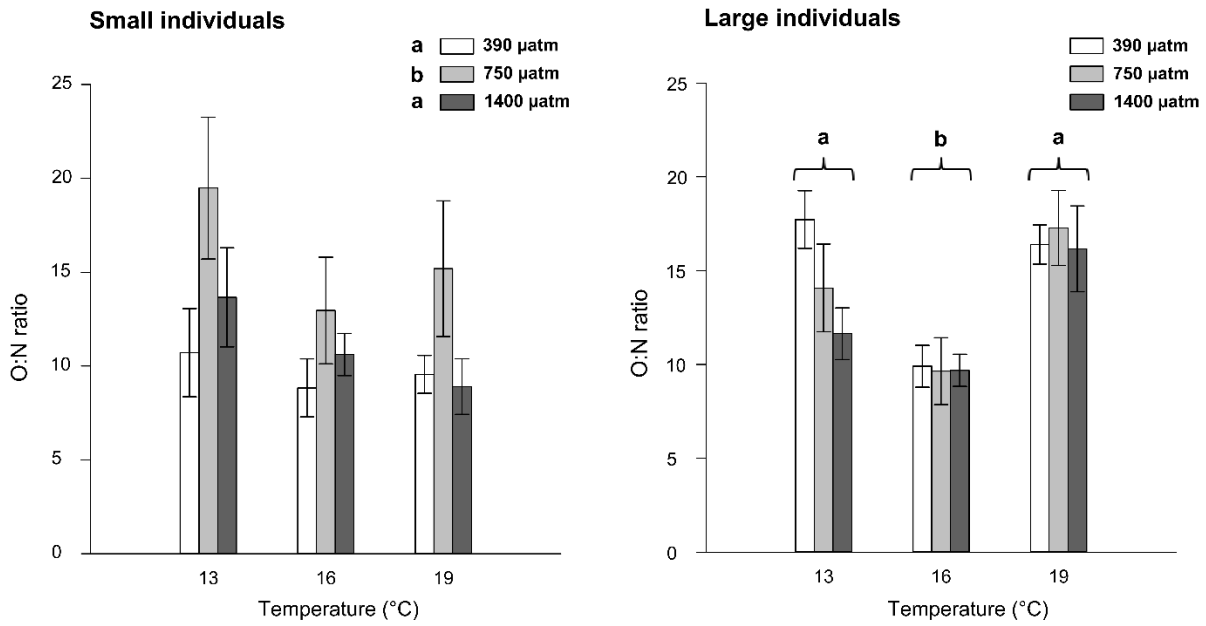
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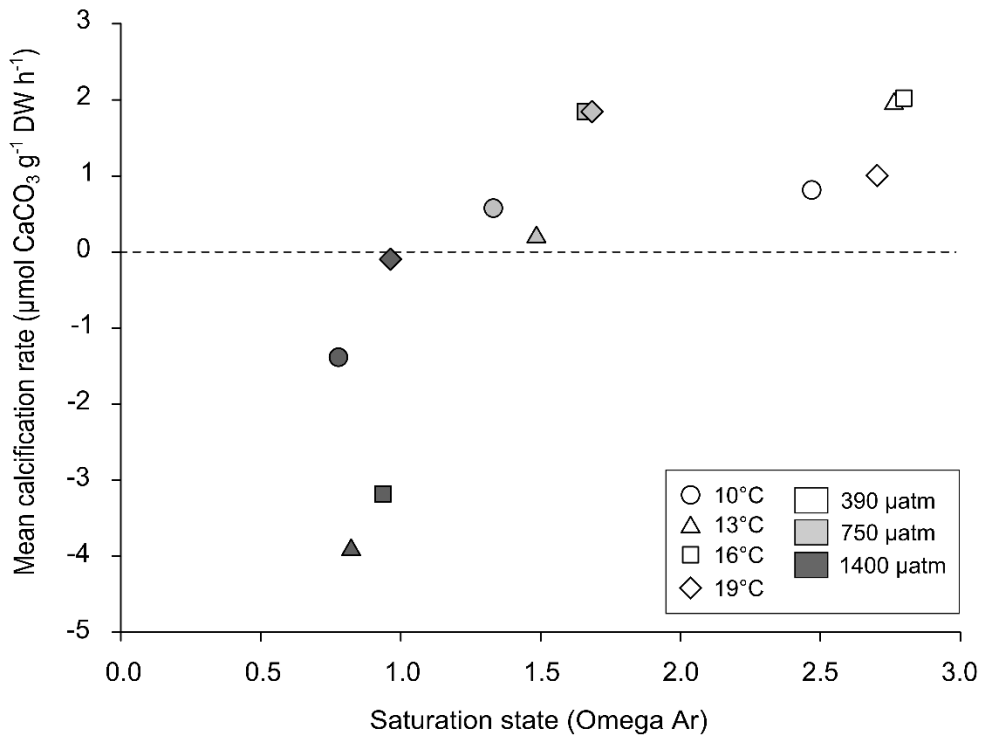
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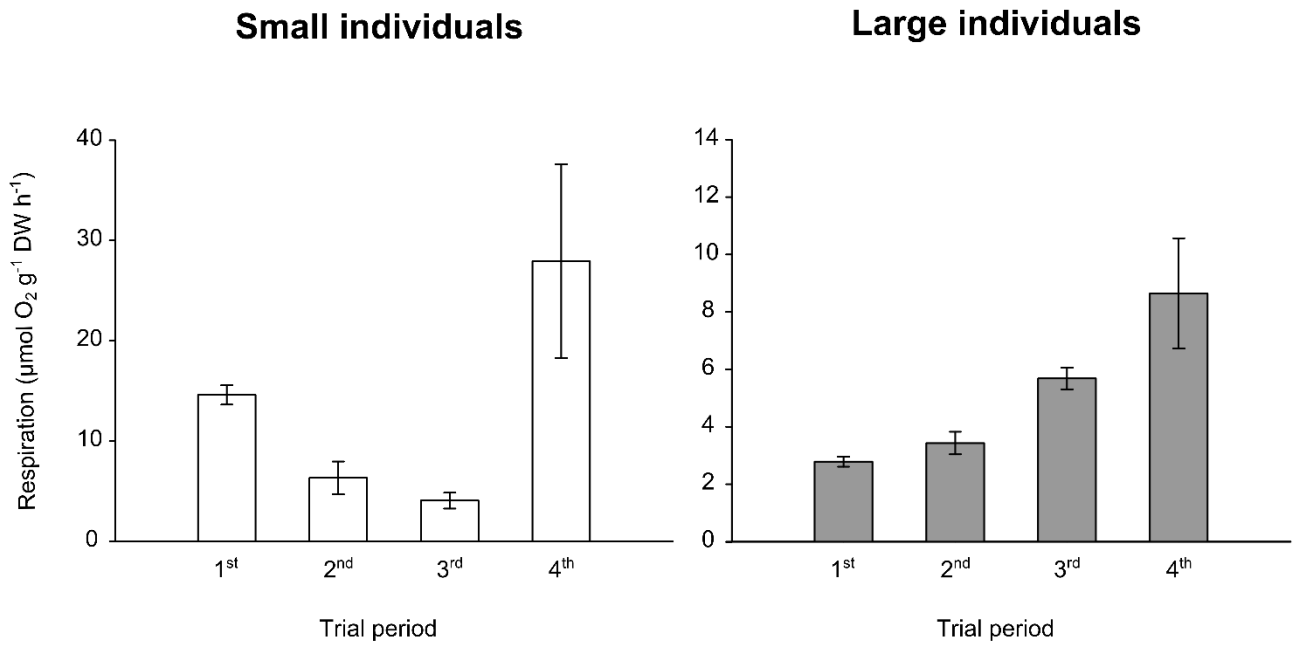
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