

1 Baseline concentrations of biliary PAH metabolites in perch (*Perca fluviatilis*)  
2 in the open Gulf of Finland and in two coastal areas

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## 22 Abstract

23 Female perch (*Perca fluviatilis*) were sampled annually between late July and early  
24 August in the eastern Gulf of Finland to monitor biliary PAH metabolite concentrations.  
25 Sampling was carried out in the open sea off Haapasaari island from 2006 to 2009 and at two  
26 coastal locations east and west of the city of Hamina in 2008. Of the PAH metabolites, only  
27 1-hydroxypyrene (1-OH pyrene) was detected at quantifiable levels in the bile of perch, and it  
28 was detected in nearly all perch. In addition, the total body weight and length and the liver  
29 and gonad weight were recorded. PAH metabolite concentrations were compared between the  
30 open sea and coastal samples and were examined in relation to body characteristics (body  
31 weight and length and proportional liver and gonad weight). There was no temporal trend in  
32 the concentration of biliary 1-OH pyrene in perch from Haapasaari. At the coastal locations,  
33 1-OH pyrene concentrations in the bile of perch were significantly higher than in the open sea  
34 Haapasaari area. Some correlations between the body characteristics of perch and 1-OH  
35 pyrene were detected when analysed separately for annual observations, but none in the whole  
36 data set. It is concluded that PAH metabolites in the bile of fish could be measured in the Gulf  
37 of Finland to detect oil spills in the open sea, and the cost-effective total fluorescence method  
38 could be used in such monitoring programmes.

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40 Keywords: PAH metabolite, perch, bile, monitoring, 1-hydroxypyrene, Baltic Sea, Gulf of  
41 Finland

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## 44 1. Introduction

45 The Baltic Sea is a highly polluted sea area due to its shallowness, long coastline and  
46 high human population with various industrial plants in the catchment area. The International  
47 Maritime Organization (IMO) has classified the Baltic Sea as a Particularly Sensitive Sea  
48 Area (PSSA) that needs special protection and into which no pollutants should be released.

49 However, vessel traffic, including oil transportation, is heavy and continually increasing  
50 (HELCOM, 2010). This is especially the case for the Gulf of Finland following the  
51 construction of large oil terminals on the Russian coast of the gulf. Both accidental and  
52 deliberate small ( $< 1 \text{ m}^3$ ) oil spills occur at a rate of hundreds per year in the whole Baltic Sea  
53 area, although the number has been decreasing (HELCOM, 2015). Fortunately, no  
54 catastrophic oil accidents have occurred in the Gulf of Finland, but quite many accidents with  
55 oil spillages  $>100 \text{ t}$  occurred in the area between 1970–1987 (Keinänen et al., 2012).

56 Oil contains polyaromatic hydrocarbons (PAHs), which are toxic to biota, including fish  
57 (Billiard et al., 2008; Tuvikene, 1995). Fish are exposed to PAHs, for example, through  
58 respiration, ingestion of contaminated food and dermal absorption, and PAHs may interfere  
59 with growth and reproduction, damage the immune system and cause lesions and tumours of  
60 the skin and liver (Logan, 2007). In the North Sea, which receives pollutants from  
61 atmospheric fallout and off-shore oil platforms, there have been indications of changes in  
62 tissue structure and altered disease frequencies in fish (Hylland et al., 2006).

63 Fish are generally able to metabolize PAHs relatively quickly, and these compounds do  
64 not therefore bioaccumulate in fish tissues (Jonsson et al., 2004; Maccubbin et al., 1988;  
65 Meador et al., 1995; Tuvikene, 1995). However, such activity causes an extra metabolic load  
66 and takes energy from other processes, such as growth. The half-lives of six PAHs (not  
67 including pyrene) were determined to be 1–4 days in rainbow trout (*Salmo gairdneri*), except  
68 for phenyl naphthalene, with a half-life of 25 days (Niimi and Dookhran, 1989). During a ten-

69 day recovery from a one-day exposure to 2 mg WSF of crude oil L<sup>-1</sup>, 1-OH pyrene  
70 concentrations in the bile of perch females decreased on average by 39% from an initial value  
71 of 2 800 ng g<sup>-1</sup> (Fahmy, 2013).

72 Recent exposure of fish to PAHs can be detected by analysing PAH metabolites in bile  
73 (Aas et al., 2000; Ariese et al., 1993; Beyer et al., 2010; Vuontisjärvi et al., 2004; Vuorinen et  
74 al., 2006). The total concentration of PAH metabolites can be analysed by measuring  
75 fluorescence at certain wavelengths (FF method), or individual PAH metabolites can be  
76 determined by gas chromatography or high performance liquid chromatography (HPLC)  
77 (Beyer et al., 2010). The measurement of biliary PAH metabolites by the either of the  
78 chromatographic techniques is very specific and exposure to other environmental toxicants  
79 does not interfere with the measurement (Beyer et al., 2010). In addition, the concentrations  
80 of biliary PAH metabolites provide a very good dose–response relationship, and if not  
81 exposed to PAHs, no metabolites are detected (Collier and Varanasi, 1991). The FF method  
82 has been suggested to be adequate for monitoring purposes (Vuontisjärvi et al., 2004).

83 The most commonly detected PAH metabolite in fish bile has been 1-hydroxypyrene  
84 (Nagel et al., 2012; Ruczynska et al., 2011; Vuontisjärvi et al., 2004). Pyrene is a four-ring  
85 PAH compound, and crude oils contain approximately 3% PAHs, i.e., 3-6-ring PAHs (Neff,  
86 1979). In the Baltic Sea environment, 1-hydroxyphenanthrene was also detected in a few  
87 flounders (*Platichthys flesus*) and in most of the eelpouts (*Zoarces viviparus*) out of seventy  
88 sampled specimens (Vuontisjärvi et al., 2004). Concentrations of PAH metabolites in fish bile  
89 are affected by gender, season and the feeding status (Brumley et al., 1998; Kammann, 2007;  
90 Richardson et al., 2004; Vuorinen et al., 2006).

91 In the Baltic Sea area, concentrations of PAH metabolites in fish bile had not been  
92 monitored over a period of years before the present study, and only sporadic investigations  
93 had been carried out (Table 1). In the EU project BEEP (Lehtonen et al., 2006), a three-year

94 sampling campaign was performed to investigate spatial, seasonal and gender effects on bile  
95 PAH metabolite concentrations in perch (*Perca fluviatilis*), flounder and eelpout (Vuorinen et  
96 al., 2006). These metabolites were also measured in the bile of perch caught in the Stockholm  
97 archipelago during a three-year survey campaign (Hansson et al., 2006b). Ruczynska et al.  
98 (2011) sampled flounder in 2008 from the Gulf of Gdansk and analysed bile for PAH  
99 metabolites. Vuorinen et al. (2003) and Pikkarainen (2006) measured PAH metabolites in the  
100 bile of perch caught in one-off sampling from the Gulf of Finland in 2001. In the North Sea,  
101 monitoring of bile PAH metabolites in dab (*Limanda limanda*) and flounder is included in the  
102 OSPAR convention (OSPAR, 2008), and the measurement of biliary PAH metabolites has  
103 been suggested to be adopted in monitoring in the Baltic Sea (Lehtonen et al., 2006).

104

Species	Sampling location	Sampling time	1-OH pyrene, ng g <sup>-1</sup>	Reference
Perch, <i>Perca fluviatilis</i>	Gulf of Finland	2001	55–160 <sup>a</sup>	Vuorinen et al. (2003)
Salmon, <i>Salmo salar</i>	Gulf of Riga	1997	140 <sup>a</sup>	Vuorinen et al. (2003)
Salmon, <i>Salmo salar</i>	Baltic Proper, SD28	1997	180 <sup>a</sup>	Vuorinen et al. (2003)
Salmon, <i>Salmo salar</i>	Åland Sea, SD29	1997	13 <sup>a</sup>	Vuorinen et al. (2003)
Perch, <i>Perca fluviatilis</i>	Western Gulf of Finland	2001	213–1149 <sup>b</sup>	Pikkarainen (2006)
Perch, <i>Perca fluviatilis</i>	Stockholm archipelago	1999–2001	20–1300 <sup>a</sup>	Hansson et al. (2006b)
Perch, <i>Perca fluviatilis</i>	Baltic Proper	2001, 2002	0–440 <sup>b</sup>	Vuontisjärvi et al. (2004)
Flounder, <i>Platichthys flesus</i>	Baltic Proper	2001, 2002	0–1000 <sup>b</sup>	Vuontisjärvi et al. (2004)
Eelpout, <i>Zoarces viviparus</i>	Baltic Proper	2001, 2002	0–1280 <sup>b</sup>	Vuontisjärvi et al. (2004)
Cod, <i>Gadus morhua</i>	Baltic Proper	2001, 2002	0–310 <sup>b</sup>	Vuontisjärvi et al. (2004)
Salmon, <i>Salmo salar</i>	Baltic Proper	1997	0–300 <sup>b</sup>	Vuontisjärvi et al. (2004)
Flounder, <i>Platichthys flesus</i>	Wismar Bay	2001, 2002	110–590 <sup>a</sup>	Vuorinen et al. (2006)
Flounder, <i>Platichthys flesus</i>	Lithuanian coast	2001	30 <sup>a</sup>	Vuorinen et al. (2006)
Flounder, <i>Platichthys flesus</i>	Denmark Strait	2004	54–92 <sup>a</sup>	Kammann (2007)
Dab, <i>Limanda limanda</i>	North Sea	2004	11–159 <sup>a</sup>	Kammann (2007)
Flounder, <i>Platichthys flesus</i>	Gulf of Gdansk	2008	20–65 <sup>a</sup>	Ruczynska et al. (2011)

106 <sup>a</sup>Range of mean concentrations; <sup>b</sup>Range of single concentration values

107           The aim of the present study was to investigate, by analysing PAH metabolites in bile  
108 samples, whether perch in the open sea archipelago at Haapasaari are exposed to PAHs and  
109 monitor the levels over a six-year period, because such monitoring has not previously been  
110 performed in the Gulf of Finland. The study was integrated with a long-term fish population  
111 status programme (Ådjers et al., 1996). In the case of a future large oil spill accident,  
112 background data on PAH metabolite concentrations would be valuable. Knowledge of the  
113 spatial variability in the PAH metabolite concentration was improved by sampling perch in  
114 two coastal bays in addition to the open sea.

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## 117 2. Materials and methods

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### 119 2.1. Sampling

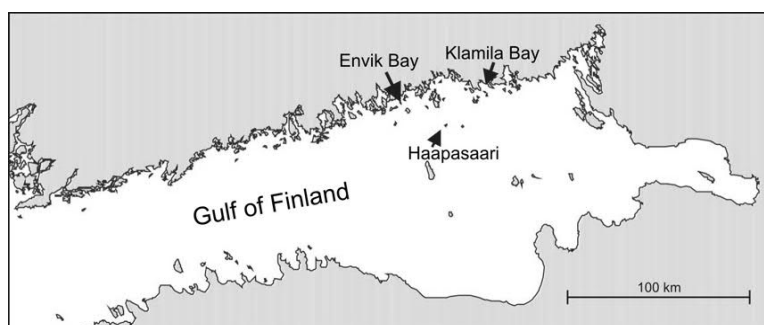
120           As part of a long-term fish monitoring programme (Ådjers et al., 1996; Rahikainen and  
121 Vähänäkki, 2006; Saulamo, 2010; Saulamo et al., 2007), perch (*Perca fluviatilis* L.) were  
122 caught by gillnets on the southern coast of Haapasaari island in the Gulf of Finland (Fig. 1)  
123 annually between the 25 July and 15 August 2005–2009, i.e., clearly after the spawning  
124 period. The knot sizes of the gillnets used to catch perch of an appropriate size for the present  
125 study were 30 and 35 mm, and the nets were set overnight for 12 hours.

126           Only female perch were selected for the present study to eliminate the random effect of  
127 sex. A bile sample was drawn into a hypodermic needle and handled as described in Vuorinen  
128 et al. (2006). Fish were weighed, the total length was measured and the liver and gonads  
129 (except in 2007) were also weighted (Table 2). A piece of liver was dissected for analysis (not  
130 reported here), and the liver and bile samples were immediately frozen in liquid nitrogen. The  
131 operculum was removed for age determination. In 2008, perch were additionally caught and

132 similarly sampled from two coastal locations, which have been planned to be possible refuge  
133 harbour areas (Fig. 1). Bile and liver samples were transported to the laboratory and bile  
134 samples were stored at  $-80\text{ }^{\circ}\text{C}$  until analysis.

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138 Fig. 1. Sampling locations for perch. Sampling was performed at Haapasaari in  
139 2005–2009 and in the coastal Envik Bay and Klamila Bay in 2008.

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142 Table 2. Mean ( $\pm$  SE) weight, length, condition factor (CF), liver somatic (LSI) and gonadosomatic (GSI) index, and the number (N) of  
 143 female perch caught in different years from the open sea at Haapasaari and two coastal locations (Envik Bay and Klamila Bay). A  
 144 different letter as a superscript denotes a significant difference ( $p < 0.05$ ) between the annual means.  
 145

Year	Location	Weight, g	Length, mm	CF	LSI	GSI	N
2005	Haapasaari	185.9 $\pm$ 7.8 <sup>b</sup>	248 $\pm$ 4 <sup>b</sup>	1.21 $\pm$ 0.04 <sup>a</sup>	1.05 $\pm$ 0.04 <sup>a</sup>	0.89 $\pm$ 0.05 <sup>a</sup>	31
2006	Haapasaari	163.4 $\pm$ 7.0 <sup>b</sup>	241 $\pm$ 3 <sup>b</sup>	1.12 $\pm$ 0.02 <sup>a</sup>	1.14 $\pm$ 0.10 <sup>a</sup>	0.91 $\pm$ 0.08 <sup>a</sup>	57
2007	Haapasaari	183.5 $\pm$ 11.6 <sup>b</sup>	245 $\pm$ 6 <sup>b</sup>	1.17 $\pm$ 0.02 <sup>a</sup>			34
2008	Haapasaari	164.3 $\pm$ 18.5 <sup>b</sup>	235 $\pm$ 9 <sup>b</sup>	1.19 $\pm$ 0.03 <sup>a</sup>	0.99 $\pm$ 0.09 <sup>a</sup>	0.95 $\pm$ 0.11 <sup>a</sup>	12
2008	Envik Bay	101.3 $\pm$ 12.7 <sup>a</sup>	203 $\pm$ 8 <sup>a</sup>	1.15 $\pm$ 0.04 <sup>a</sup>	0.94 $\pm$ 0.10 <sup>a</sup>	0.92 $\pm$ 0.10 <sup>a</sup>	12
2008	Klamila Bay	97.3 $\pm$ 12.5 <sup>a</sup>	204 $\pm$ 8 <sup>a</sup>	1.07 $\pm$ 0.03 <sup>a</sup>	0.91 $\pm$ 0.05 <sup>a</sup>	0.74 $\pm$ 0.05 <sup>a</sup>	12
2009	Haapasaari	174.3 $\pm$ 11.5 <sup>b</sup>	243 $\pm$ 4 <sup>b</sup>	1.18 $\pm$ 0.04 <sup>a</sup>	1.13 $\pm$ 0.05 <sup>a</sup>	0.95 $\pm$ 0.08 <sup>a</sup>	26

147

## 148 2.2. Chemical analysis

149 Biliary PAH metabolites were analysed by HPLC as described in Vuontisjärvi et al.  
150 (2004), with changes in volumes as described in the following. Bile samples (10  $\mu\text{L}$ ) were  
151 hydrolyzed with  $\beta$ -glucuronidase / aryl sulfatase (10  $\mu\text{L}$ , Merck) at 37 °C for 2 hours in 150  
152  $\mu\text{L}$  (total volume topped with Millipore purified  $\text{H}_2\text{O}$ ). Proteins were precipitated with HPLC-  
153 grade acetonitrile (150  $\mu\text{L}$ ) and the samples were centrifuged (Heraeus Fresco 21) at 10 000  
154 rpm for 5 minutes. The supernatant was filtered through a 0.2- $\mu\text{m}$  syringe filter and a 5- $\mu\text{L}$   
155 aliquot was injected into a UHPLC system with a fluorescence detector (excitation and  
156 emission wavelengths 346 nm and 384 nm, respectively). An external standard curve was  
157 prepared from 2 to 100 ng 1-hydroxypyrene (1-OH pyrene)  $\text{mL}^{-1}$  and it was always run with  
158 bile sample sets. The uncertainty of analysis was 20% and the quantification limit 5.0  $\text{ng g}^{-1}$ .  
159 The recovery was 93% and results were corrected for this recovery.

160 The chromatographic system consisted of a Waters Acquity UPLC binary pump,  
161 autosampler and fluorescence detector. The separation column comprised a C-18 type BEH  
162 UPLC column (1.7  $\mu\text{m}$  particles, 1 x 100 mm) and the pump was operated at a flow rate of  
163 150  $\mu\text{L min}^{-1}$ . Separation was performed with a gradient program from 10% to 80%  
164 acetonitrile in the B pump and with 0.1% trifluoroacetic acid in water in the A pump. The  
165 retention time of 1-OH pyrene was 6.5 minutes and the total run time was 15 minutes.

166 As a quality control, a laboratory control sample prepared from fish bile was run with  
167 each sample set. The laboratory has participated in intercalibration exercises for the  
168 determination of PAH metabolites in fish bile samples both by HPLC and FF (Kammann et  
169 al., 2013).

170

## 171 2.3. Statistical calculations

172 The condition factor (CF) of perch was calculated as  $CF = 10^5 * \text{weight (g)} / \text{length}$   
173  $(\text{mm})^3$  and the relative liver (LSI) and gonad (GSI) weight as a percentage of the body weight.

174 Variance homogeneity was tested using Levene's test. The statistical significance of  
175 differences between the years and locations was tested by one-way ANOVA with the *post hoc*  
176 Student-Neuman-Keuls test for significant differences ( $p < 0.05$ ) between the sample means.

177 Statistical tests were performed using the Statistical Analysis System, SAS ver 9.4 (SAS  
178 Institute Inc., 2008).

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### 180 3. Results and discussion

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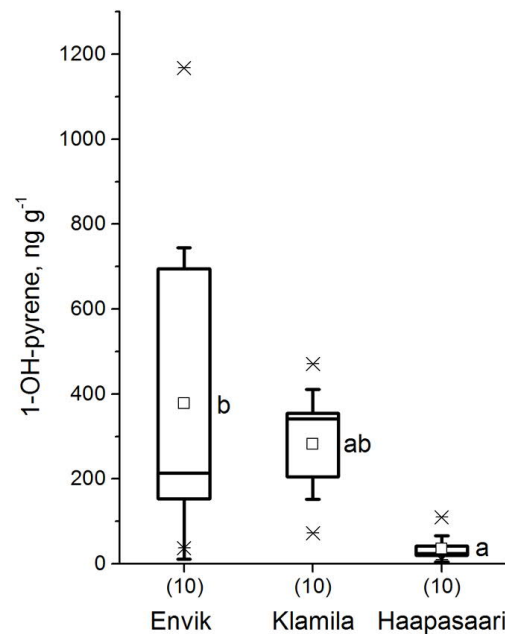
#### 182 3.1. Coastal versus open sea locality

183 Of the 16 analysed PAH metabolites, only 1-OH pyrene was detected at quantifiable  
184 levels in the bile of perch. The concentrations of 1-OH pyrene in perch bile were several times  
185 higher at the two coastal locations, Envik Bay and Klamila Bay, than in the open sea  
186 archipelago of Haapasaari in the same year, and significantly ( $p < 0.05$ ) higher in Envik Bay  
187 (Fig. 2). The mean concentrations at the two coastal locations did not differ significantly from  
188 each other. Anthropogenic sources are probably reflected in the bile of the coastal perch,  
189 because in addition to oil spills, 1-OH pyrene may originate from various burning processes  
190 and also be transported via the atmosphere (Anderson and Lee, 2006), and apparently also  
191 with run-off waters. This is supported by the fact that the mean 1-OH pyrene concentration  
192 and its variation were numerically higher at Envik Bay, which is located closer to an  
193 industrial and highly populated area than the Klamila Bay. As reviewed by Tuvikene (1995),  
194 PAHs entering the aquatic environment are mostly localised in rivers, estuaries and coastal  
195 waters. The coastal mean values of the present study were up to two to three times higher than  
196 have earlier been measured in the bile of perch caught near an oil refinery at Sköldvik, Gulf of

197 Finland, in 2001 (Vuorinen et al., 2003). In the bile of perch caught in 2001 from the western  
 198 coast of the Gulf of Finland (Table 1), 1-OH pyrene concentrations analysed by HPLC were  
 199 even higher (Pikkarainen, 2006), being nearly twice as high as in the present study in perch  
 200 from Envik Bay. At highly contaminated sites, the 1-OH pyrene concentrations in fish bile  
 201 may be much higher than in the perch of the present study. For example, 2000–300 000 ng 1-  
 202 OH pyrene g<sup>-1</sup> bile was detected in English sole (*Parophrys vetulus*) caught from the side of  
 203 the Puget Sound polluted by aromatic hydrocarbons (Krahn et al., 1987).

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207 Fig. 2. Box plots of 1-OH pyrene concentrations in the bile of female perch caught in two  
 208 coastal bays (Envik Bay and Klamila Bay) and in the open sea at Haapasaari in  
 209 2008. Whiskers depict 1 x SD, asterisks the minimum and maximum  
 210 observations, upper and lower parts of the boxes 25 and 75% of observations,  
 211 horizontal lines the median value and small squares the mean concentrations. A  
 212 different letter close to the means denotes a significant difference ( $p < 0.05$ )  
 213 between the means. The sample size is indicated in parentheses.

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216 *3.2. Temporal variation in biliary PAH metabolites of perch*

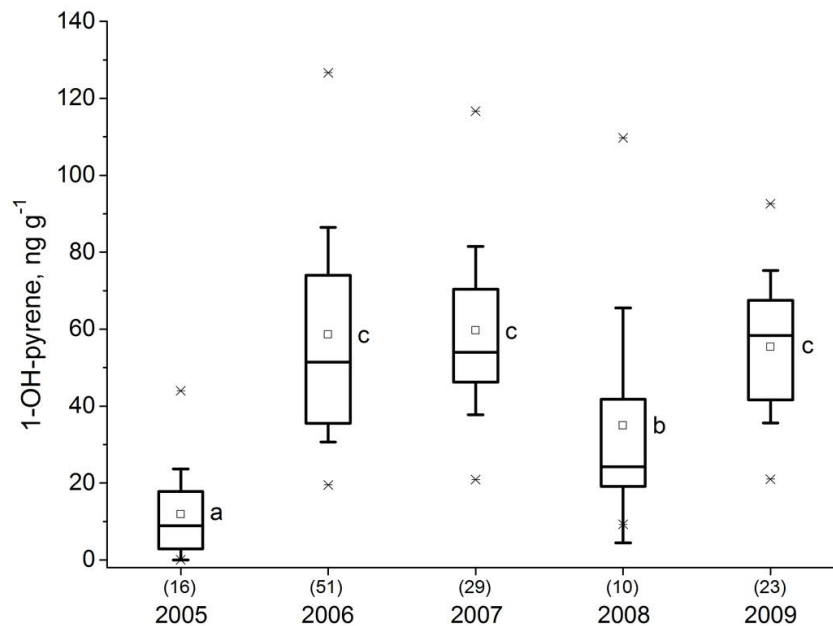
217 Biliary 1-OH pyrene concentrations of perch from the open sea archipelago off  
218 Haapasaari differed significantly ( $p < 0.05$ ) between years, the concentrations having been  
219 clearly and significantly lowest in the first monitoring year, 2005 (Fig. 3). However, no clear  
220 temporal trend was observed in perch during the five-year monitoring period. Indeed, the  
221 mean 1-OH pyrene concentrations in perch bile in 2006, 2007 and 2009 were nearly equal,  
222 with a significant decrease in 2008 (Fig. 3). In all these years, the 1-OH pyrene concentrations  
223 were on average clearly lower than in the bile of perch in the two coastal bays. The variation  
224 among fish at all sampling times was large (Fig. 3), although less than in the coastal locations.  
225 However, 1-OH pyrene was detected in nearly all samples. This indicates that perch are  
226 continuously exposed to pyrene, but individual differences in behaviour or feeding status may  
227 cause the variation in pyrene concentrations. Oil is sparingly dissolved in water, and there  
228 may have been large differences in exposure of individual fish. In laboratory experiments, 1-  
229 OH pyrene concentrations were also quite variable in the bile of perch exposed to the water  
230 soluble fraction (WSF) of crude oil (Fahmy, 2013).

231 On average, the 1-OH pyrene concentrations in the bile of perch were approximately  
232 one half of those detected in the bile of salmon caught from the open sea of the Baltic Proper  
233 or the Gulf of Bothnia in 1997 (Vuorinen et al., 2003). In flounder caught in the Baltic Sea  
234 near the Denmark Strait in 2004, the biliary concentrations of 1-OH pyrene were on average  
235 similar to those in perch of the present study, but those in dab from the North Sea were on  
236 average lower than in perch (Kammann, 2007). Ruczynska et al. (2011) detected similar  
237 concentrations of 1-OH pyrene in the bile of flounders from the Gulf of Gdansk (Table 1) to  
238 those in perch of the present study. The low concentrations of biliary 1-OH pyrene in perch  
239 also reported in other studies (Table 1) suggest that pollution by PAH compounds is  
240 continuous. No major oil accidents have occurred in the Baltic Sea during the 2000s, but

241 100–140 ship accidents, mostly in harbours, have been reported annually (Keinänen et al.,  
 242 2012), in addition to open sea spills (HELCOM, 2015). As 1-OH pyrene also originates from  
 243 various burning processes, it has apparently been transported via the atmosphere (Anderson  
 244 and Lee, 2006).

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249 Fig. 3. Box plots of 1-OH pyrene concentrations in the bile of perch caught in the open  
 250 sea at Haapasaari in 2005–2009. Whiskers depict 1 x SD, asterisks the minimum  
 251 and maximum observations, upper and lower parts of the boxes 25 and 75% of  
 252 observations, horizontal lines the median value and small squares the mean  
 253 concentrations. A different letter close to the means denotes a significant  
 254 difference ( $p < 0.05$ ) between the means. The sample size is indicated in  
 255 parentheses.

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257

258 The lack of a temporal trend may be due to effective control measures by authorities.

259 All HELCOM member states regularly conduct aerial surveillance flights to monitor illegal

260 and accidental oil spills in the Baltic Sea (HELCOM, 2010), and the number and size of these  
261 has decreased from about 600 oil spills in the early 1990s to close 100 in 2014 (HELCOM,  
262 2015). This development has occurred despite the fact that oil transportation in the Baltic Sea  
263 more than doubled between 1997–2008 (HELCOM, 2010) and was predicted to increase in  
264 the Gulf of Finland by more than ten times from 1995 to 2015 (Jolma and Hietala, 2009); in  
265 fact, oil transportation increased by a factor of seven (web reference). It is possible that  
266 exposure of perch to PAHs has actually decreased along with the diminished number of oil  
267 spills reported (HELCOM, 2015). This could be an explanation for the lower levels of 1-OH  
268 pyrene in the bile of perch of the present study compared to those in samples of feeding  
269 salmon from the open sea in autumn ten years earlier (Vuorinen et al., 2003).

270

### 271 *3.3. Biliary PAH metabolites and body characteristics of perch*

272 Perch caught from the Haapasaari waters in different years did not differ from each  
273 other in terms of mean body length or weight, but those caught at the two coastal locations  
274 were significantly ( $p < 0.05$ ) smaller than the perch from Haapasaari (Table 2). When  
275 analysing the whole data set, the average CF, LSI and GSI (Table 2) were similar ( $p > 0.05$ )  
276 in all perch populations. However, taking weight as a covariate, the average CF of perch from  
277 Haapasaari was significantly ( $p < 0.05$ ) higher than that of perch from Klamila Bay,  
278 potentially reflecting better feeding resources in the open sea area. The CF of perch from  
279 Envik Bay did not differ significantly from the CF of perch of the two other areas in 2008. In  
280 that year, the GSI of perch from Klamila Bay was significantly ( $p < 0.05$ ) lower than that of  
281 perch from Envik Bay, but the GSI of perch from Haapasaari did not differ ( $p > 0.05$ ) from  
282 those of the coastal locations. Long-term monitoring data on perch from Swedish reference  
283 areas in the Baltic Proper revealed a decrease in the GSI that along with changes in some  
284 other variables was interpreted to result from environmental pollution (Hansson et al., 2006a).

285           The concentrations of 1-OH pyrene in bile did not correlate with length, weight or CF,  
286 or with the proportional liver (LSI) or gonad (GSI) weight of female perch from Haapasaari in  
287 all data combined (Table 3). However, calculating Pearson correlations separately for each  
288 monitoring year revealed significant negative correlations between 1-OH pyrene and weight ( $r$   
289 = -0.310,  $p = 0.027$ ) and CF ( $r = -0.290$ ,  $p = 0.039$ ) in 2006, when the highest individual 1-OH  
290 pyrene concentrations were detected. Moreover, there were nearly significant negative  
291 correlations between the bile 1-OH pyrene concentration and the LSI in 2005 ( $r = -0.492$ ,  $p =$   
292  $0.053$ ) and in 2009 ( $r = -0.389$ ,  $p = 0.055$ ). In a laboratory experiment, exposure of fasting  
293 perch for 7 and 21 days to the water-soluble fraction of crude oil resulted in a decrease in the  
294 LSI, and significant negative correlations of the LSI with 1-OH pyrene, 2-OH naphthalene  
295 and 1,2-OH chrysene in females, and with the two last mentioned metabolites in males  
296 (Vuorinen et al., 2003). The proportional weight of gonads (GSI) correlated positively and  
297 highly significantly ( $p < 0.0001$ ) with the proportional liver weight (LSI). This apparently  
298 results from the liver synthesising vitellogenin to be transported via blood to the developing  
299 ovaries. Consequently, the liver weight increases during vitellogenesis (Mommensen and  
300 Walsh, 1988). This development thus appears rather normal in the early stage of the  
301 reproductive cycle, despite perch being exposed, evidently slightly, to the PAH compound  
302 pyrene. On the other hand, higher PAH metabolite levels in bile of eelpout from the Gulf of  
303 Finland compared to the Gulf of Riga were associated with a lower LSI and condition factor  
304 and with higher geno- and cytotoxicity and parasite infections (Kreitsberg et al., 2012).  
305 Moreover, pollution from offshore oil platforms in the North Sea has been suspected to  
306 associate with structural changes in tissues and both decreases and increases in some disease  
307 prevalences in fish (Hylland et al., 2006).  
308  
309



310 Table 3. Pearson correlation coefficients with the p-value and number of observations (in  
 311 parentheses) in female perch caught from Haapasaari waters in 2005–2009.  
 312

	Length	CF	LSI	GSI	1-OH pyrene
Weight	0.912	0.499	-0.105	-0.145	-0.134
	<.0001	<.0001	0.240	0.106	0.124
	(160)	(160)	(126)	(125)	(132)
Length		0.204	-0.234	-0.231	-0.126
		0.010	0.008	0.010	0.150
		(160)	(126)	(125)	(132)
CF			-0.121	-0.137	-0.123
			0.179	0.129	0.160
			(126)	(125)	(132)
LSI				0.705	-0.098
				<.0001	0.325
				(125)	(102)
GSI					-0.077
					0.442
					(102)

313

314

#### 315 3.4. Biliary PAH metabolites as a biomarker

316 PAH metabolites in the bile of fish provide a specific and dose-responsive indication of  
 317 exposure to PAHs (Collier and Varanasi, 1991) or a biomarker of ongoing or recent exposure  
 318 of fish to PAHs (Beyer et al., 2010; Lehtonen et al., 2006). An extensive oil spill at the  
 319 Butinge oil terminal on the Lithuanian coast was still detectable by the FF method as bile  
 320 PAH metabolites in flounders after seven months (Barsiene et al., 2006). This biomarker has  
 321 been used in many studies to demonstrate accidental exposure of fish to PAHs (Kammann,  
 322 2007; Kreitsberg et al., 2010; Lee and Anderson, 2005; Nagel et al., 2012; Ruddock et al.,

2002; Ruddock et al., 2003), and included as one of the biomarkers in the monitoring programme of OSPAR in the North Sea (OSPAR, 2008). Thus far, unequivocal evidence linking this specific biomarker with higher order effects such as reproductive failures is lacking (Lee and Anderson, 2005), as is the case for biomarkers in general (Forbes et al., 2006). However, the combination of biomarkers with advanced statistical analysis, such as principal component analysis, might reveal associations between the physiological status of fish and environmental contaminants (Gagnon and Rawson, 2016). Biomarkers might also be related to phenotyping responses in fish (Houde et al., 2014), as was the approach in the present study.

Variability was detected in the biliary 1-OH pyrene concentrations of individual perch in the present study. While no clear pattern was detected, this might, along with variation in exposure, partly result from differences in the feeding status of perch on various sampling occasions. Samples of the present study were collected annually at the same time and only from female perch, and seasonal and sex effects were thus excluded (Kammann, 2007; Ruddock et al., 2002; Vuorinen et al., 2006). Bile is a part of the normal digestion, and the bile volume as well as the concentration of metabolites in bile are therefore affected to some degree by the feeding status of fish (Brumley et al., 1998; Collier and Varanasi, 1991; Richardson et al., 2004).

Attempts have been made to minimize random error in biliary PAH metabolite concentrations when sampling populations by standardisation procedures, such as normalising the results against the bile biliverdin or protein concentration (Collier and Varanasi, 1991; Kammann, 2007; Ruddock et al., 2002; Ruddock et al., 2003; Vuorinen et al., 2006). In the present study, such procedures were not tested, although biliverdin normalisation has been reported to reduce variation (Collier and Varanasi, 1991; Kammann, 2007; Ruddock et al., 2003). However, biliverdin or protein normalisation, although it affected metabolite results

348 measured by the FF method, did not affect results for 1-OH pyrene in the HPLC method  
349 (Vuorinen et al., 2006) that was used in the present study. Moreover, biliverdin normalisation  
350 did not consistently reduce variation in various PAH metabolite concentrations measured in  
351 the bile of plaice (*Pleuronectes platessa*), and thus it was suggested that PAH metabolite  
352 concentrations should be preferably reported both as raw data and normalised by biliverdin  
353 (Richardson et al., 2004). Measurement of total fluorescence is simple and cost effective, as  
354 bile samples only need to be diluted before measuring fluorescence. Because no other PAH  
355 metabolites except 1-OH pyrene were detected in perch bile in the present study, and total  
356 fluorescence correlates well with the bile 1-OH pyrene concentration, FF would be the  
357 method of choice for monitoring purposes (Vuontisjärvi et al., 2004; Vuorinen et al., 2006).

358

#### 359 4. Conclusions

360

361 On the basis of PAH metabolite concentrations in the bile of perch caught in the open  
362 sea archipelago, pollution by PAH compounds is continuous in the Gulf of Finland. However,  
363 the biliary PAH metabolite concentrations in perch are rather low in the open sea and several  
364 times lower than at the coastal locations. Therefore, larger changes in PAH exposure at open  
365 sea locations should be readily detectable. Thus, the present monitoring data on PAH  
366 metabolites in fish bile in the Gulf of Finland provide a background reference in case of a  
367 larger oil accident. Because no other PAH metabolites except 1-OH pyrene were detected in  
368 bile, the cost-effective FF method would be appropriate for monitoring purposes.

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