



This is a repository copy of *Accelerated carbonation of reactive MgO and Portland cement blends under flowing CO₂ gas.*

White Rose Research Online URL for this paper:
<http://eprints.whiterose.ac.uk/157828/>

Version: Accepted Version

Article:

Wang, L., Chen, L., Provis, J.L. orcid.org/0000-0003-3372-8922 et al. (2 more authors) (2020) Accelerated carbonation of reactive MgO and Portland cement blends under flowing CO₂ gas. *Cement and Concrete Composites*, 106. 103489. ISSN 0958-9465

<https://doi.org/10.1016/j.cemconcomp.2019.103489>

Article available under the terms of the CC-BY-NC-ND licence
(<https://creativecommons.org/licenses/by-nc-nd/4.0/>).

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: <https://creativecommons.org/licenses/>

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk
<https://eprints.whiterose.ac.uk/>

1 Accelerated Carbonation of Reactive MgO and Portland Cement Blends 2 Under Flowing CO₂ Gas

3 Lei Wang^{1,2}, Liang Chen¹, John L. Provis², Daniel C.W. Tsang^{1,*}, Chi Sun Poon¹

4

5 ¹ Department of Civil and Environmental Engineering, The Hong Kong Polytechnic University, Hung Hom,
6 Kowloon, Hong Kong, China.

7 ² Department of Materials Science and Engineering, The University of Sheffield, Sir Robert Hadfield
8 Building, Mappin St, Sheffield S1 3JD, United Kingdom.

9 * Corresponding author: dan.tsang@polyu.edu.hk

10

11 **Abstract**

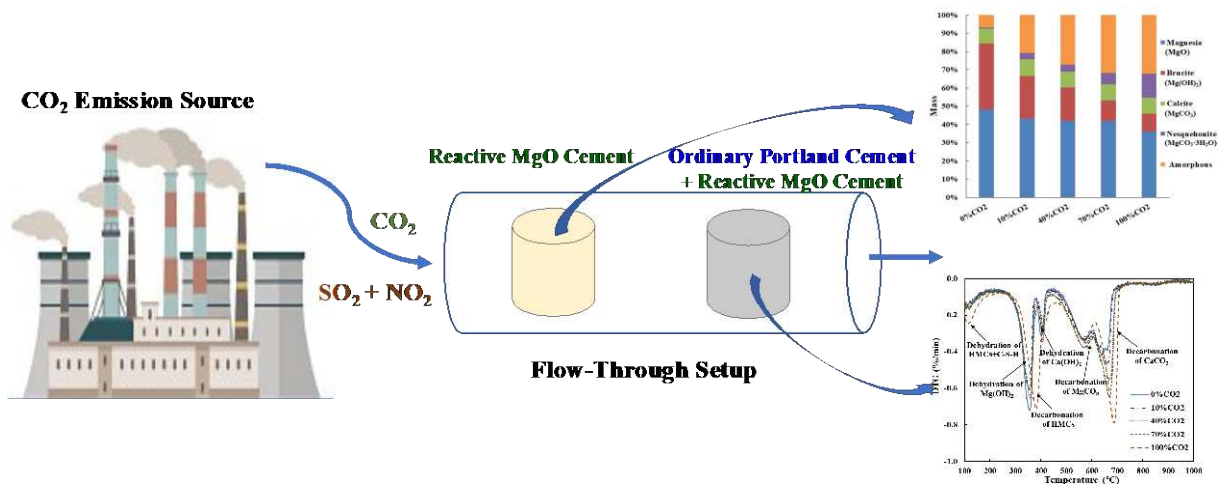
12 The use of MgO-based materials for sequestration of CO₂ offers technical advantages and
13 environmental incentives. However, the understanding of accelerated carbonation of MgO-
14 based materials in flowing CO₂ gas is limited. This study elucidates the carbonation behaviour
15 of reactive MgO cement (MC) and MgO-Portland binary cement (BC) in a simulated CO₂-rich
16 industrial exhaust. Quantitative X-ray diffraction and thermogravimetric analyses showed that
17 nesquehonite (MgCO₃·3H₂O) was the major carbonation product in MC pastes, whereas
18 CaCO₃ was preferentially generated in BC pastes. The relative humidity of exhaust gas
19 influenced CO₂ diffusion and the carbonation rate; 98% humidity facilitated MC carbonation
20 whereas 50% was favourable for BC carbonation. Although CO₂ concentration determined the
21 carbonation rate, 10% CO₂ gas in the exhaust was sufficient to accelerate carbonation. The
22 carbonation degree and compressive strength of samples cured for 7 days at 10% CO₂ were
23 comparable to the values of samples cured for 1 day at 100% CO₂. The presence of acid gases
24 during CO₂ curing inhibited the carbonation and hydration processes, but the presence of
25 Portland cement in the BC system gave good compatibility with acids and relieved the
26 inhibitory effect. Desulphurization and denitrification of industrial exhaust gas are nonetheless

27 desirable before CO₂ curing. This study builds the foundation for utilising industrial CO₂
 28 exhaust to accelerate the carbonation of Mg-based materials.

29 **Keywords:** eco-friendly cement; CO₂ sequestration/utilisation; amorphous hydrated carbonate;
 30 cement hydration chemistry; gaseous waste valorisation; sustainable chemistry/engineering.

31

32 **Graphical Abstract**



33

34 **Highlights:**

- 35 • MgO-based cement can sequester and utilize CO₂ from industrial exhaust.
- 36 • Relative humidity influenced CO₂ diffusion and carbonation rate.
- 37 • 7-d curing with 10% CO₂ concentration ensured sufficient carbonation degree.
- 38 • Acid gases in exhaust severely inhibited carbonation and hydration.
- 39 • Binary MgO-Portland cement showed fast carbonation rate and good compatibility with
 40 acids.

41

42 **1. Introduction**

43 Cement and concrete have been essential materials for the urbanisation of society. More than
 44 10 billion tonnes of concrete are annually generated worldwide, which has been the second-
 45 most-used material in the world only behind water [1]. Concrete is usually based on ordinary

46 Portland cement (PC) as a binding material, due to its robust, reliable, and inexpensive nature.
47 PC is generated from calcination of limestone and clay minerals at 1400 °C in a kiln, which is
48 associated with a high CO₂ emission (660–820 kg CO₂ per tonne) contributing to
49 approximately 7-10% of global anthropogenic CO₂ emissions [2-4]. Therefore, extensive
50 research has been devoted to developing alternative low-carbon cementitious systems to
51 partially or totally replace PC in certain applications.

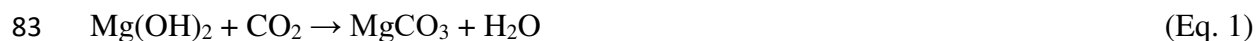
52

53 Reactive magnesia cement (MC) is a promising candidate “green cement”, produced by the
54 calcination of magnesite at relatively low temperature (650–800 °C) [5,6]. Recently, recovery
55 of Mg²⁺ from salt lakes and seawater to produce MC has been proposed as an eco-friendly
56 approach [7], as the low geological abundance of high-purity magnesium minerals in many
57 parts of the world constrains the uptake of magnesia-based cements [8]. However, in locations
58 where suitable resources are available at scale, this is a potentially attractive route to
59 valorisation of under-utilised resources. For instance, as a by-product of the Li₂CO₃ production
60 from salt lakes, there are 200,000 tonnes of MgO produced annually in Qarhan Salt Lake in
61 Qinghai Province, China [9]. Partial MC replacement in PC-based cementitious binder systems
62 could enhance the properties of cement and concrete products, including characteristics such
63 as denser pore structure, higher mechanical strength, superior resistance to chemical attack and
64 corrosion, as well as excellent compatibility with contaminants [10,11].

65

66 Recent research has shown that MC is a versatile material that can be used alone as a binder,
67 or blended with other materials for diverse applications, such as MgO expansive cements,
68 MgO-based self-healing materials, MgO-based stabilization/solidification agents, MgO-
69 modified alkali-activated materials, magnesium phosphate cements, magnesium oxychloride
70 (Sorel) cements, magnesium oxysulphate cements, magnesium silicate hydrate (M-S-H)

71 cements, and others [8,12,13]. Importantly, MC derived from an efficiently-produced magnesia
72 source is regarded by some as a “carbon negative” cementitious material, as it can sequester
73 CO₂ and gain improved binding properties during accelerated carbonation [14,15]. CO₂ curing
74 has been suggested to accelerate the carbonation of cement-based products and facilitate carbon
75 sequestration in solid mineral phases [16,17]. During CO₂ curing, dissolved and ionised CO₂
76 induces the carbonation of Ca²⁺/Mg²⁺ ions from the cement matrix, which then precipitates in
77 the voids of the matrix as carbonates (anhydrous and hydrated) in a short period of time,
78 boosting setting and hardening process, forming a dense, strong and potentially stable structure
79 [18-20]. MC usually contains more than 85% active MgO, and the CO₂ sequestration capacity
80 can reach up to 92.8 wt%, which is higher than the capacity of PC (50.4 wt%) based on
81 theoretical calculation. The carbonation of MC can take place via the formation of magnesium
82 carbonate (MgCO₃) from hydrated Mg(OH)₂ by the uptake of CO₂:



84 or with the incorporation of water to form hydrated magnesium carbonates (HMCs), including
85 nesquehonite (MgCO₃·3H₂O), dypingite (Mg₅(CO₃)₄(OH)₂·5H₂O), and artinite
86 (Mg₂(OH)₂CO₃·3H₂O). In most cases, nesquehonite is the most prominent Mg carbonate phase
87 during CO₂ curing:



89 HMCs can form a well-densified structure with good binding ability, thus enhancing the
90 mechanical properties of MC-based products [8,21]. The reaction pathways and quantitative
91 compositions of HMCs are kinetically and thermodynamically controlled by many factors such
92 as temperature, CO₂ pressure, ionic strength, etc [22, 23].

93

94 Accelerated carbonation of MC is well documented in the recent literature [24,25]. However,
95 most of these studies have reported accelerated carbonation that was performed in a static CO₂

96 curing chamber with pressured and concentrated CO₂ gas [26,27]. This static CO₂ curing can
97 make the maximum use of CO₂ for accelerated carbonation and facilitate the investigation of
98 the mechanisms of MC carbonation. However, some researchers have pointed out that the static
99 CO₂ curing represents an ideal CO₂ curing, which is an energy-intensive approach and far from
100 practical application [28]. As the carbonation of MC is rate-limited and quite slow in an
101 atmospheric environment, it has been argued that the re-adsorption of CO₂ from the
102 environment will not take place to a meaningful extent in MC during its service life, meaning
103 that the classification of “carbon negative” is doubtful [8]. The carbonation of MC and PC
104 binary system has attracted wide attention, because the binary cement could utilise their
105 respective advantages and alleviate individual drawbacks, achieving satisfactory carbonation
106 efficiency, early strength, and durability [27,29]. Up to now, there is no consensus that MC and
107 MC-based cementitious materials can be effectively carbonated to sequester CO₂ in practical
108 situations.

109

110 In this study, we investigate the potential efficacy of accelerated carbonation of MC and MgO-
111 based binary cement (BC) in a field-relevant situation. The flow-through CO₂ curing system
112 can simulate a real process exhaust situation, which allows adjustable temperature, humidity,
113 flow rate, and CO₂ concentration. The flow-through curing system can utilise CO₂-rich
114 industrial exhaust gas as a CO₂ source for carbonation, e.g., flue gases from fossil-fuel power
115 plants, cement rotary kilns, steelworks, and refuse incineration plants, which may contain CO₂
116 concentrations ranging from 5% to 20% [30]. Because the carbonation rate is sensitive to
117 exposure conditions during CO₂ curing [31], the effects of parameters in the exhaust gas supply
118 (CO₂ concentration, relative humidity) on carbonation of MC and BC are investigated.
119 Additionally, exhaust CO₂ gas is usually associated with acid gases (such as SO₂ and NO₂),
120 which should be considered before practical applications, so these contaminants are included

121 in the gas environment for selected experiments.

122

123 To explore the carbonation efficiency of MgO-based cement in simulated industrial exhaust
124 gases, this study aims to: (i) assess the effect of relative humidity on the CO₂ diffusion and
125 carbonation of MgO pastes; (ii) elucidate the influence of CO₂ concentration on the carbonation
126 rate and formation of final carbonates; (iii) evaluate the interference of acid gases with the
127 accelerated carbonation of MgO systems; and (iv) investigate the interactions between MC and
128 PC in the flow-through CO₂ curing system.

129

130 **2. Materials and Methodology**

131 *2.1 Materials and Sample Preparation*

132 Reactive MgO cement and ASTM Type I ordinary Portland cement were used in this study.
133 The MC with a density of 3.15 g cm⁻³, produced from the calcination of MgCO₃ at 700 °C, was
134 obtained from Renheng Magnesium Company, Liaoning Province, China. The PC with a
135 density of 3.16 g cm⁻³ was purchased from Green Island Cement Limited, Hong Kong. The
136 chemical compositions of the PC and MC cement were determined by X-ray fluorescence
137 (XRF) and are listed in Table 1. The particle size distribution and the X-ray diffraction (XRD)
138 data obtained for PC and MC are presented in Figure S1 and Figure S2 (Supplementary
139 Information). Reagent-grade Mg(NO₃)₂, NaCl, and K₂SO₄ used for adjusting the relative
140 humidity were purchased from Tianjin Chemical Reagent Factory, China. Reagent-grade
141 H₂SO₄ (95-98%) and HNO₃ (70%) solution for simulating acid gases were purchased from
142 Sigma-Aldrich.

143

144 The MC binder and a binary cement (BC) binder were investigated in the flow-through CO₂
145 curing condition. The BC binder was composed of 50 wt% MC and 50 wt% PC. The water-to-

146 binder ratio was kept at 0.25 in both cases, giving zero slump paste mixtures, which was
147 advantageous for block production and subsequent CO₂ curing [29]. To simulate the
148 interference of acid gases (SO₂ and NO₂), H₂SO₄ (0.125 wt% of paste) and HNO₃ (0.0137 wt%
149 of paste) were added into specified mixtures based on theoretical calculation (in Supplementary
150 Information). For the production of MC and BC blocks, water was added into the binder and
151 mixed for 3 min by a planetary stirrer. The fresh pastes were cast in steel cylindrical moulds
152 (55 mm internal diameter) and compacted with 30 MPa pressure until the height of the sample
153 was compressed to 55 mm. After an additional 1 min compaction, the samples were demoulded
154 immediately without slump and subjected to different curing conditions. In the flow-through
155 CO₂ curing chamber, the relative humidity of gas was controlled at 50%, 75%, and 98% by
156 equilibration with saturated Mg(NO₃)₂, NaCl, and K₂SO₄ solutions respectively; the CO₂
157 concentration was adjusted to 0%, 10%, 40%, 70%, and 100%; the curing times (i.e., periods
158 of CO₂ curing) were 6 h, 1 d, 3 d, and 7 d. All experiments were conducted in triplicate and the
159 average values are reported with error bars where appropriate. The CO₂ concentration was
160 adjusted by controlling the blending ratio of variable flow rates of air and CO₂ gas, and a CO₂
161 meter (CM-0003, CO₂Meter) was used to monitor and record the CO₂ concentration in the
162 flowing gas. A schematic diagram of the flow-through CO₂ curing set up is shown in **Figure**
163 **1**.

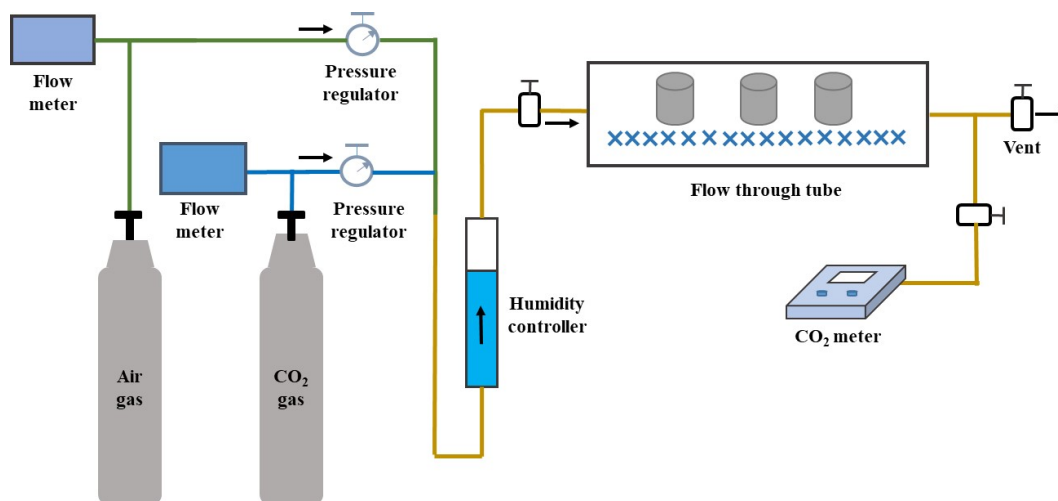
164

165 **Table 1.** Chemical compositions of PC and MC (wt %)

| | Na ₂ O | MgO | Al ₂ O ₃ | SiO ₂ | P ₂ O ₅ | SO ₃ | K ₂ O | CaO | Fe ₂ O ₃ | LOI |
|-----------|-------------------|-------|--------------------------------|------------------|-------------------------------|-----------------|------------------|-------|--------------------------------|------|
| <i>PC</i> | 0.17 | 1.38 | 4.79 | 18.99 | 0.08 | 4.52 | 0.79 | 65.72 | 3.10 | 2.45 |
| <i>MC</i> | 0.00 | 90.46 | 0.46 | 5.81 | 0.10 | 0.10 | 0.15 | 2.22 | 0.59 | 5.42 |

166 LOI: loss on ignition; PC: ordinary Portland cement; MC: reactive MgO cement.

167



168

169 **Figure 1.** Schematic diagram of flow-through CO₂ curing setup.

170

171 **2.2 Physical Properties and Thermogravimetric/Spectroscopic/Microscopic Analyses**

172 The uniaxial compressive strength of the MC and BC blocks was examined by using a standard
 173 testing machine (Testometric CXM 500-50 KN) at a loading rate of 0.6 MPa s⁻¹ [32].

174 Thermogravimetric analysis (TGA) of powder samples was performed by heating from 40°C
 175 to 1000 °C at 10 °C min⁻¹ with argon purge gas (Rigaku Thermo Plus). Crystalline-phase

176 mineralogy of samples was determined using a high-resolution powdered X-ray diffractometer
 177 (XRD, Rigaku SmartLab) in the range 2θ = 15-45° at a rate of 3° min⁻¹, with Cu Kα adiation

178 generated at 45 kV and 200 mA. It should be noted that the XRD scan at a low angle is also
 179 important for detecting the strong peaks of hydrated magnesium carbonates in this range. For

180 quantitative X-ray diffraction (Q-XRD) analysis, 20 wt% corundum (Al₂O₃) was used as an
 181 internal standard to determine the content of amorphous constituents, although it was regarded

182 as a semi-quantitative method in view of the likelihood of errors. The Q-XRD analysis was
 183 calculated by the whole powder pattern fitting (WPPF) method of the integrated X-ray powder

184 diffraction software (PDXL). The surface morphology and elemental distribution of samples
 185 with Au coating were investigated by scanning electron microscopy with energy-dispersive X-

186 ray spectroscopy (SEM-EDX, TESCAN VEGA3 XM) at an accelerating voltage of 15 kV with

187 a current of 70-78 μA . For quality assurance, all of the samples were crushed into specified
188 size and homogeneously mixed for the analytical tests.

189

190 **3. Results and Discussion**

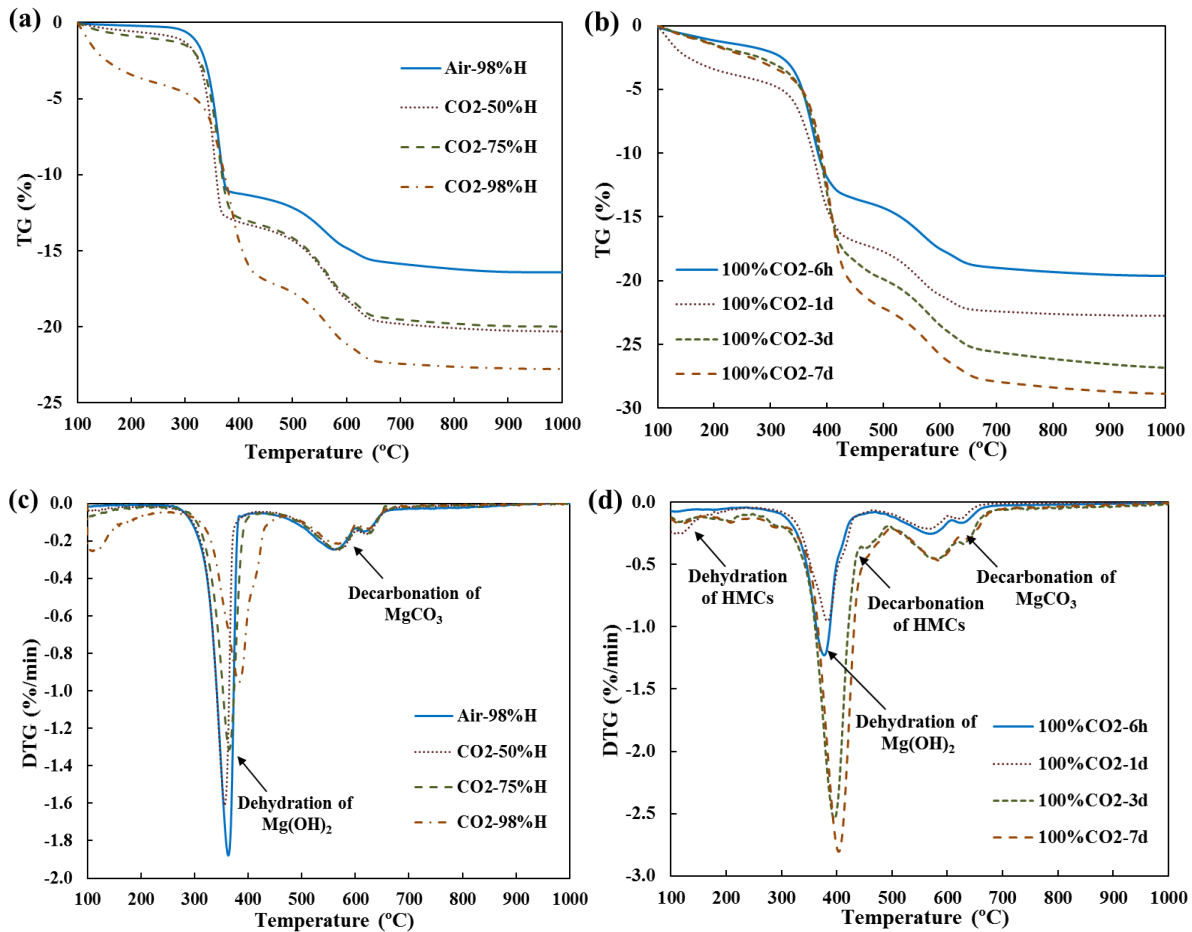
191 *3.1 The Role of Relative Humidity in Hydration and Carbonation*

192 *3.1.1 MC Pastes*

193 From the thermogravimetry (TG) curves (Figure 2a), the weight loss from 100 °C to 300 °C
194 can be ascribed to the evaporation of the physically adsorbed water and bound water from the
195 hydrated magnesium carbonates (HMCs, e.g. nesquehonite) [27]. The sharp peaks near 370 °C
196 and 390 °C in the derivative thermogravimetry (DTG) curves (Figure 2b) correspond to the
197 dehydroxylation of $\text{Mg}(\text{OH})_2$ and decarbonisation of HMCs, respectively [27]. The mass drop
198 at higher temperatures (from 500 °C to 600 °C) was associated with the decomposition of
199 metastable Mg-carbonates. The peaks at temperatures above 600 °C mainly result from the
200 decomposition of stable, well-crystallized MgCO_3 [23]. In the 1-d air cured samples, there was
201 a mass loss of 12.3 wt% at 370 °C (Figure 2a), equivalent to 39.7 wt% $\text{Mg}(\text{OH})_2$ content. The
202 air cured samples also showed 4.32 wt% mass loss due to carbonates (equivalent to 8.25 wt%
203 MgCO_3), which resulted from the original MgCO_3 in raw MgO cement (Figure S2). By
204 comparison, after 1-d CO_2 curing, the $\text{Mg}(\text{OH})_2$ peak was reduced and shifted from 370 °C to
205 390 °C, especially in the samples with 98% humidity CO_2 curing. An obvious dehydration peak
206 of HMCs could be observed in the 98% humidity samples. Thus, 98% humidity was selected
207 as the optimal humidity for the carbonation of MC samples, although 50% humidity was
208 regarded as an optimal humidity for PC system in a previous study [16]. This is attributed to
209 the relatively high water consumption for the dissolution of $\text{Mg}(\text{OH})_2$ ($K_{\text{sp}} = 1.13 \times 10^{-11}$ at 25
210 °C) compared to the $\text{Ca}(\text{OH})_2$ ($K_{\text{sp}} = 5.02 \times 10^{-6}$ at 25 °C) resulting from Portland cement
211 hydration [33]. It should be noted that the content of MgCO_3 (approximately 10.32%) was

212 similar in all 1-d samples. This suggests that $Mg(OH)_2$ was preferentially carbonated into
 213 HMCs [34], because the ΔG of $MgCO_3 \cdot 3H_2O$ generation ($-38.7 \text{ kJ mol}^{-1}$) was lower than the
 214 energy of $MgCO_3$ generation ($-30.2 \text{ kJ mol}^{-1}$) (Figure S3).

215



216

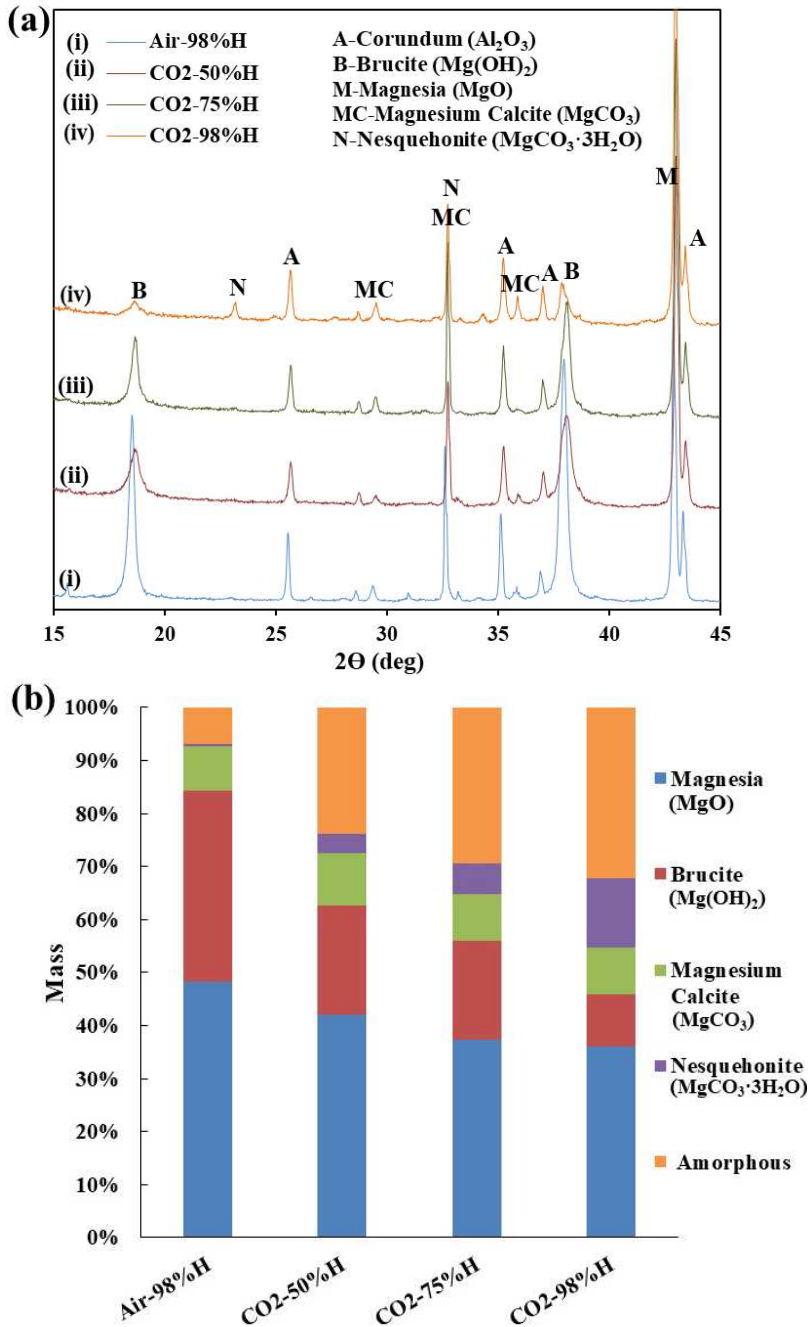
217

218 **Figure 2.** TGA of MC pastes: (a) TG curves of 1-d cured pastes as a function of curing
 219 environment and humidity; (b) TG curves of pastes at cured at 100% CO₂ and 98% humidity
 220 for different durations; (c) DTG curves corresponding to (a); (d) DTG curves corresponding to
 221 (b). (HMCs: hydrated magnesium carbonates).

222

223 The degree of carbonation in MC samples as a function of CO₂ curing time (at 98% humidity)
 224 is shown in Figure 2b&d. The content of newly formed HMCs in 6-h CO₂ samples was
 225 negligible, reflecting the low carbonation degree (i.e., the degree of the transformation of
 226 Ca(OH)₂ and Mg(OH)₂ into CaCO₃, MgCO₃ and HMCs) in the early term. However, after 1-d
 227 CO₂ curing, the peak at 370 °C was shifted to a higher temperature. The mass loss of HMCs

228 reached 11.0 wt% and 14.2 wt% in 3-d and 7-d CO₂ cured samples, respectively. Therefore,
 229 sufficient curing time is essential for continued carbonation under the flow-through curing
 230 situation where both MgO hydration and CO₂ diffusion are time-dependent processes.
 231



232

233

234 **Figure 3.** X-ray diffractograms (a) and Q-XRD analysis (b) of MC pastes cured for 1 day under

235 air and CO₂ curing, with various relative humidities as marked.

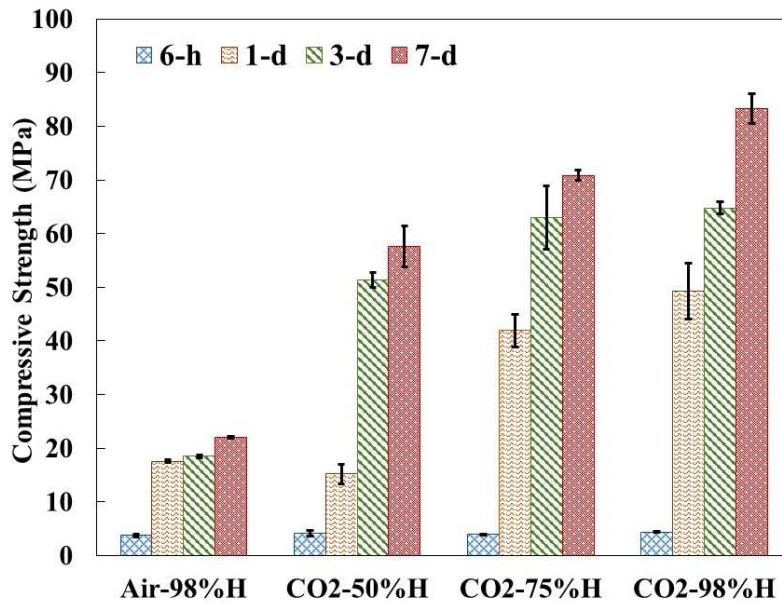
236

237 The XRD diffractograms (Figure 3a) illustrate that a remarkable peak of brucite ($\text{Mg}(\text{OH})_2$) at
238 $18.7^\circ 2\theta$ existed in 1-d air cured samples. The Q-XRD results (Figure 3b) further indicate that
239 there was 36.1 wt% of $\text{Mg}(\text{OH})_2$ content and 48.2 wt% of unreacted MgO in the air cured
240 samples. After 1-d CO_2 curing, the contents of both $\text{Mg}(\text{OH})_2$ and MgO decreased whereas the
241 contents of MgCO_3 , nesquehonite ($\text{MgCO}_3 \cdot 3\text{H}_2\text{O}$), and amorphous phase (e.g., poorly-
242 crystalline $\text{Mg}(\text{OH})_2$, MgCO_3 , and HMCs) increased, which agreed with the TGA results. In
243 particular, for the samples with 98% humidity CO_2 curing, a new peak due to nesquehonite
244 appeared at $23.2^\circ 2\theta$. The XRD results provide further evidence that high relative humidity
245 (98%) is a favourable condition for CO_2 curing of MC-based materials under flow-through
246 conditions.

247

248 Figure 4 shows that MC blocks, after 6-h air curing, had a low compressive strength (3.74
249 MPa), whereas the strength reached 17.6 MPa after 1-d air curing, and gradually increased with
250 curing time due to continued MgO hydration. In the CO_2 cured samples, there was marginal
251 enhancement of strength after 6-h CO_2 curing compared to the air-cured samples. This proved
252 that 6-h curing was insufficient for CO_2 diffusion and further carbonation under the non-
253 pressurised flow-through conditions, although accelerated carbonation has previously been
254 shown to be nearly complete under higher-pressure CO_2 within 2-h of curing [34]. After 1-d
255 CO_2 curing, the compressive strength of MC samples significantly increased, and after 7-d CO_2
256 curing at 98% humidity, it was 3.8 times higher than that of the 1-d air cured samples. These
257 results demonstrate that MC pastes can be carbonated to generate abundant HMCs, as indicated
258 by the TGA and Q-XRD results (Figure 2 and 3), for strength enhancement even in the non-
259 pressurised flow-through curing system.

260



261

262 **Figure 4.** Compressive strength of MC pastes under air and CO₂ curing, with various relative
 263 humidities as marked.

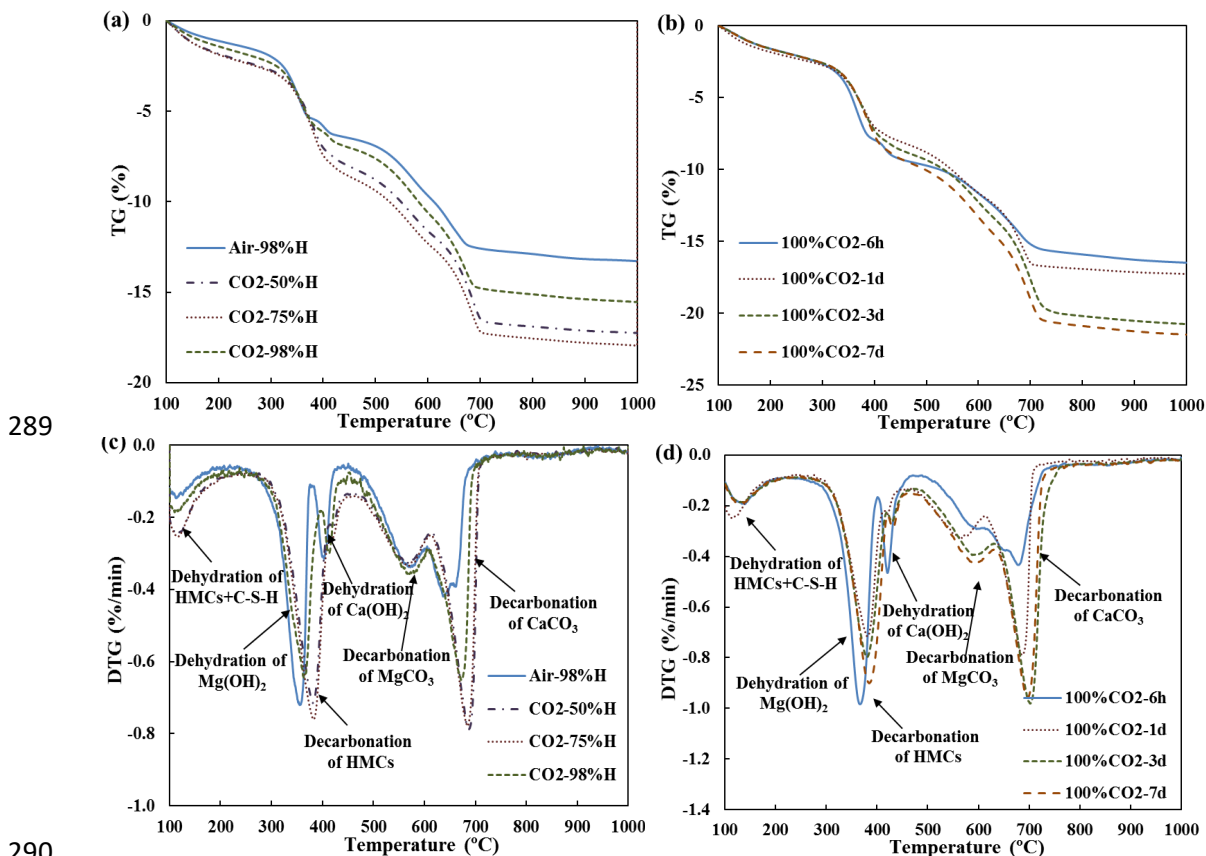
264

265 3.1.2 BC Pastes

266 The TG and DTG curves of the 1-d cured BC pastes are illustrated in Figure 5a&c. From Figure
 267 5c, 3.46 wt.% mass loss at 370 °C (equivalent to 11.7 wt.% Mg(OH)₂) and 1.04 wt.% mass loss
 268 at 410 °C (equivalent to 4.27 wt.% Ca(OH)₂) were observed in the 1-d air cured samples. By
 269 comparison, the Ca(OH)₂ peaks diminished and CaCO₃ peaks intensified after 1-d CO₂ curing,
 270 while Mg(OH)₂ was partially transformed into HMCs. Among the different relative humidity
 271 conditions tested, low humidity levels (50% and 75%) were favourable for the carbonation of
 272 BC pastes, which was distinct from the MC systems. In the BC system, the generation and
 273 dissolution of Ca(OH)₂ was much faster than Mg(OH)₂, thus, the CO₂ diffusion rate became
 274 the major limiting factor determining the overall rate of carbonation. As the CO₂ diffusion
 275 coefficient in water is 4 to 5 orders-of-magnitude lower than in air [35], low humidity levels
 276 (50% and 75%) facilitated CO₂ diffusion and increased the overall carbonation rate of the BC
 277 pastes.

278

279 Under 50% humidity, 6-h CO₂ curing significantly promoted the transformation from Ca(OH)₂
 280 to CaCO₃, however, it had little effect on the carbonation of Mg(OH)₂. This reflects that CO₂
 281 preferentially reacted with Ca to generate CaCO₃. One of the major reasons is that the ΔG of
 282 CaCO₃ generation (-73.0 kJ mol⁻¹) is lower than that of MgCO₃ generation (-30.2 kJ mol⁻¹)
 283 [34]. In addition, the dissolution and hydration of tricalcium silicate (C₃S) was much faster
 284 than that of reactive MgO in the BC system, thus the Ca(OH)₂ content was higher than the
 285 Mg(OH)₂ content at early age. After the consumption of Ca(OH)₂, Mg(OH)₂ was partially
 286 carbonated into HMCs after 1-d CO₂ curing, and then Mg(OH)₂ and HMCs were gradually
 287 transformed into MgCO₃, yet the content of hydrates and carbonates were similar in the 3-d
 288 and 7-d CO₂ cured samples.



291 **Figure 5.** TGA of BC pastes: (a) TG curves of 1-d cured pastes as a function of curing
 292 environment and humidity; (b) TG curves of pastes at cured at 100% CO₂ and 50% humidity
 293 for different durations; (c) DTG curves corresponding to (a); (d) DTG curves corresponding to
 294 (b). (BC: binary cement; C-S-H: calcium silicate hydrate; HMCs: hydrated magnesium
 295 carbonates).

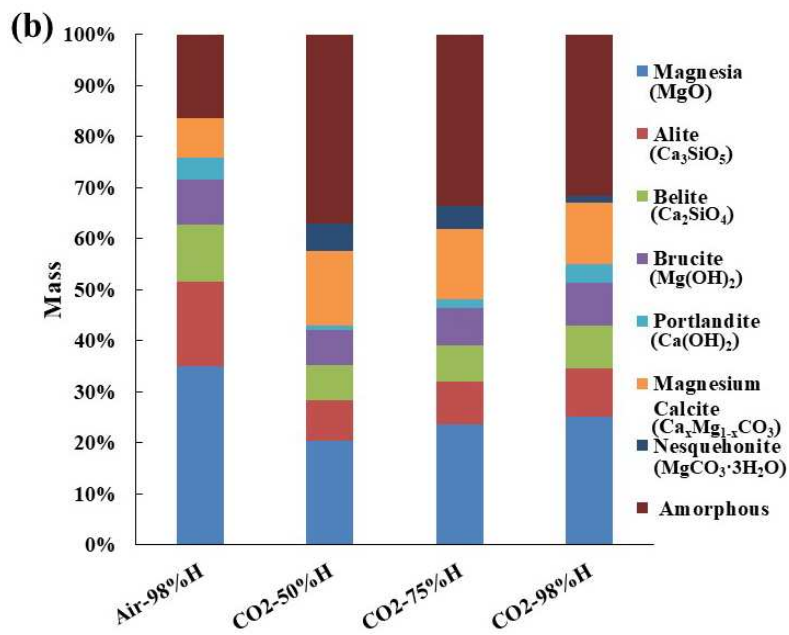
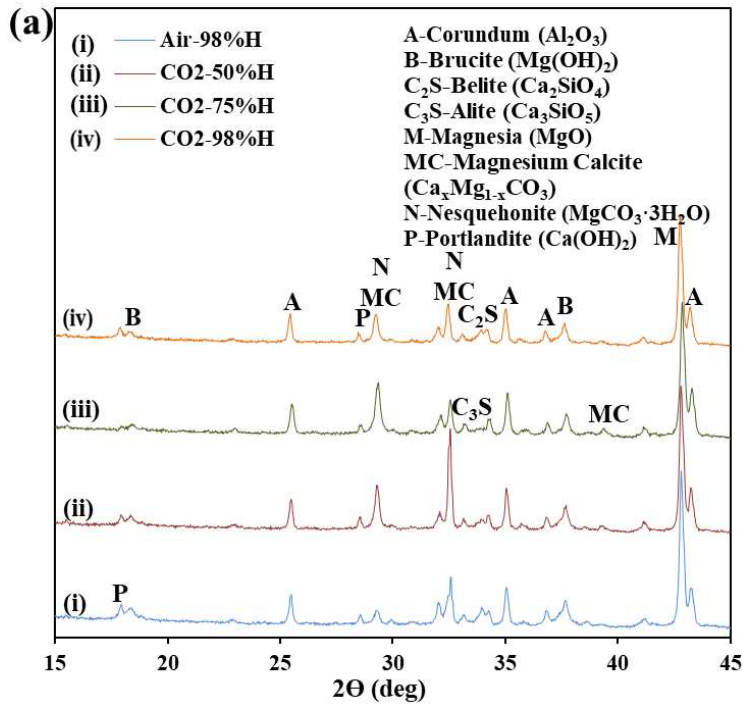
296

297 The Q-XRD results (Figure 6b) show that there were large quantities of unreacted MgO (34.8
298 wt%), C₃S (16.5 wt%), and C₂S (11.2 wt%) in the 1-d air cured samples. The 1-d CO₂ curing
299 effectively boosted the hydration and carbonation of raw materials to generate carbonates (e.g.,
300 CaCO₃). The Q-XRD results confirm that Ca(OH)₂ was preferentially carbonated during 1-d
301 CO₂ curing. In the early stage (1-d), the BC samples with 50% humidity CO₂ curing showed
302 the highest carbonation degree.

303

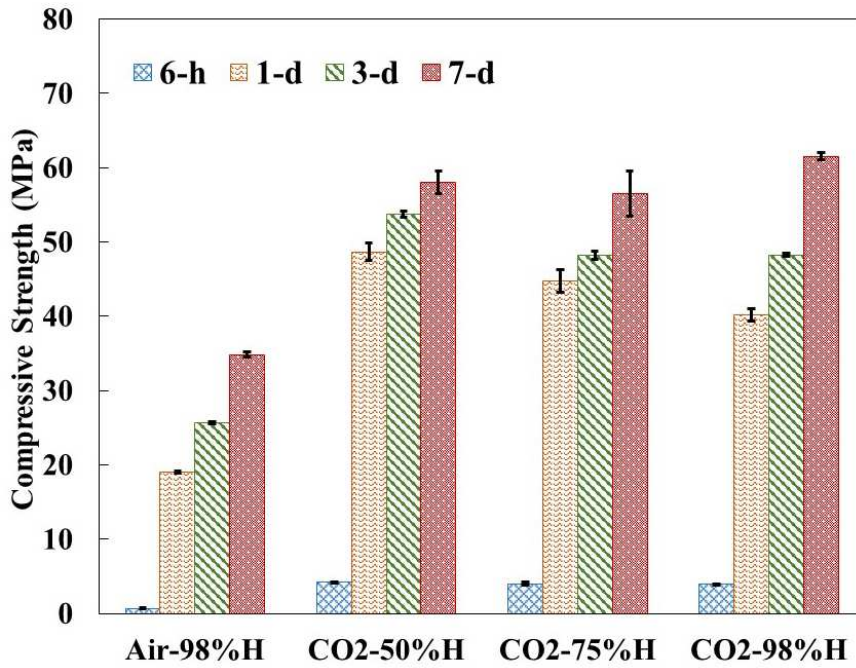
304 As shown in Figure 7, the compressive strength of BC blocks under 6-h air curing was only
305 0.67 MPa, while 6-h CO₂ curing enhanced the strength to 4.03 MPa. The carbonation rate in
306 the BC system was faster than in MC system due to the higher activity of Ca²⁺ in BC. After 1-
307 d CO₂ curing, all the carbonated samples were stronger than the 7-d air cured samples. The
308 strength of BC samples reached as high as 61.5 MPa after 7-d CO₂ curing. Among the samples,
309 50% humidity facilitated the early strength development of BC samples by accelerated
310 carbonation whereas 98% humidity promoted the later stage (7-d) strength development by
311 continued hydration and carbonation. Humidity as a key factor determining the rate of
312 carbonation can be adjusted at different stages of curing under the flow-through conditions
313 tested here, providing an interesting route to practical process optimisation.

314



317 **Figure 6.** XRD diffractograms (a) and Q-XRD analysis (b) of 1-d cured BC pastes with
 318 different humidities

319



320
 321 **Figure 7.** Compressive strengths of BC pastes under air and CO₂ curing with various
 322 humidities.

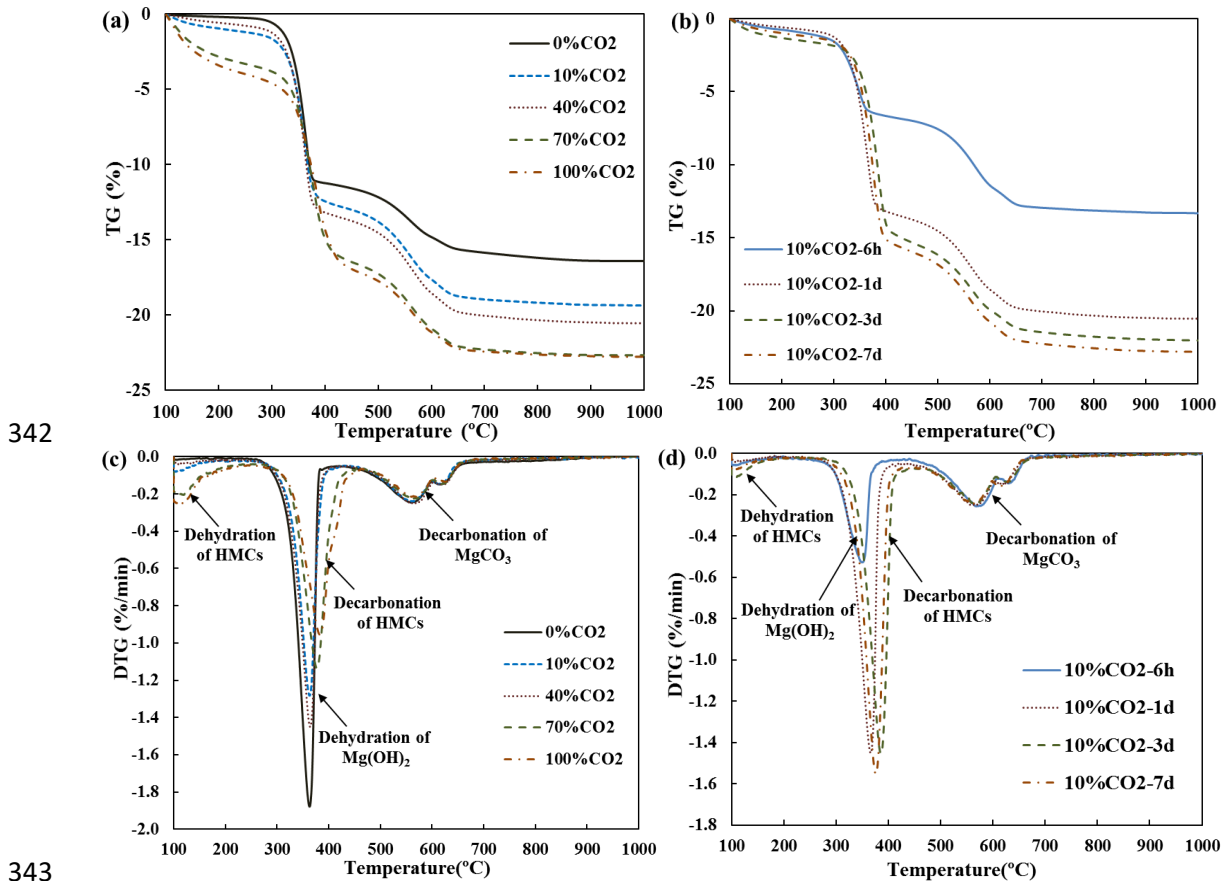
323

324 *3.2 Efficiency of Various CO₂ Concentrations in Accelerated Carbonation*

325 *3.2.1 MC Pastes*

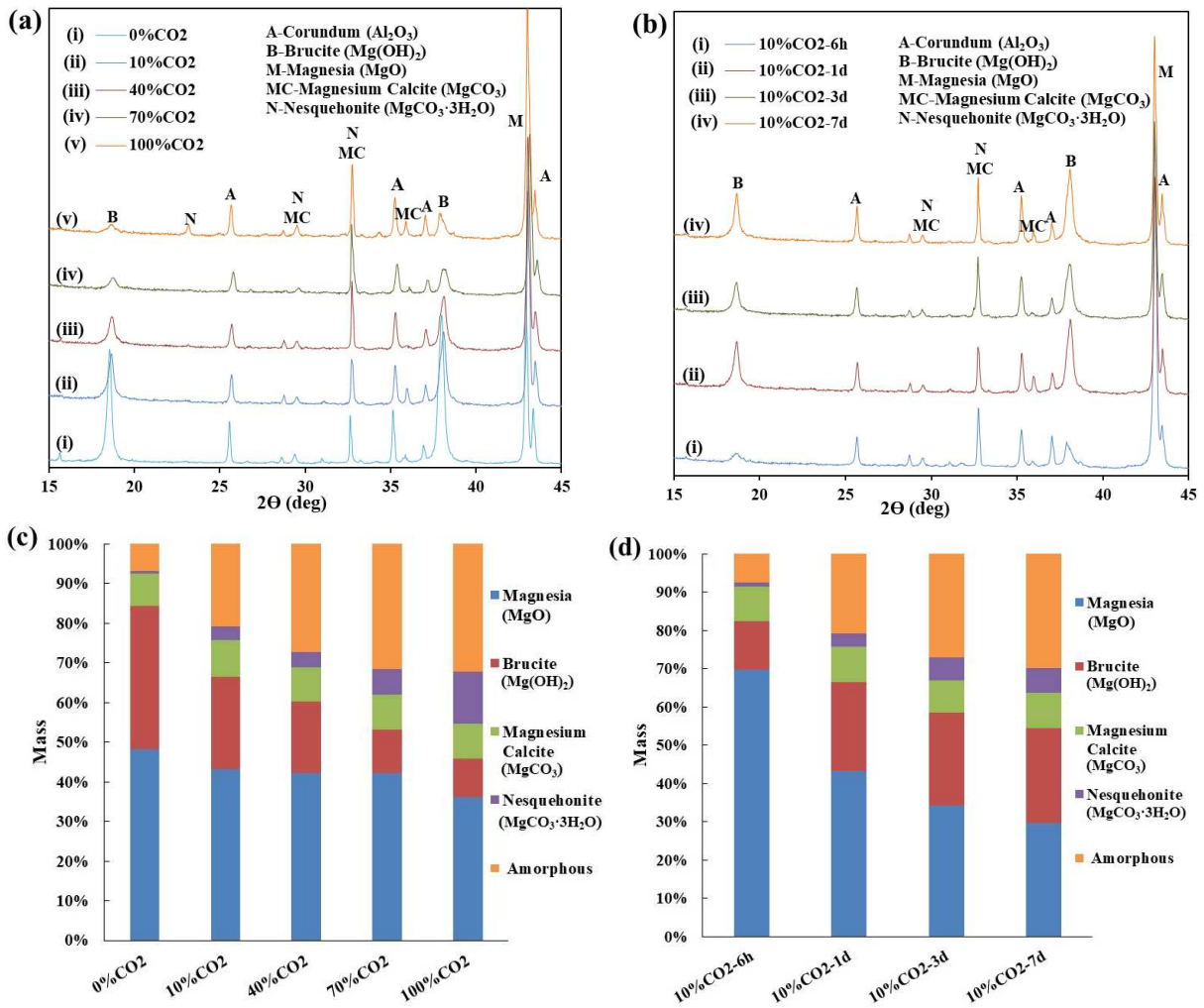
326 As illustrated by the measured TG curves (Figure 8), the mass loss of all the carbonated samples
 327 was larger than the values for corresponding air cured samples (0% CO₂). After 1-d CO₂ curing,
 328 samples that had been subjected to different CO₂ concentrations showed similar weight drops
 329 in TG due to decarbonation at 570 °C (5.15 wt%) and 630 °C (1.30 wt%). However, compared
 330 to air cured samples, the peak at 370 °C was enlarged and shifted to 390 °C, especially for 70%
 331 CO₂ and 100% CO₂ samples. The broad peak between 100 °C and 300 °C also increased in
 332 70% CO₂ and 100% CO₂ samples. These results suggest that the transformation of Mg(OH)₂
 333 into HMCs was insignificant in the low CO₂ concentration (10% and 40%) samples while it
 334 was remarkable in the high CO₂ concentration (70% and 100%) samples. The degree of
 335 carbonation in 10% CO₂ samples increased with curing time (Figure 8b). Compared to 1-d CO₂
 336 curing, the mass loss peak shifted from 370 °C to 390 °C after 3-d CO₂ curing (Figure 8d). The
 337 total weight drop of 7-d 10% CO₂ cured samples (22.8 wt%) was comparable to that of 1-d

338 100% CO₂ cured samples (22.7 wt%) (Figure 8b). Therefore, a relatively low concentration of
 339 CO₂ (e.g., 10%) under the flow-through curing conditions could still accelerate the carbonation
 340 of MC samples, and extended CO₂ curing effectively enhanced the carbonation degree.
 341



344 **Figure 8.** TGA of MC pastes: (a) TG curves of 1-d pastes as a function of CO₂ concentrations;
 345 (b) TG curves of pastes at cured at 10% CO₂ for different durations; (c) DTG curves
 346 corresponding to (a); (d) DTG curves corresponding to (b).

347



348

349

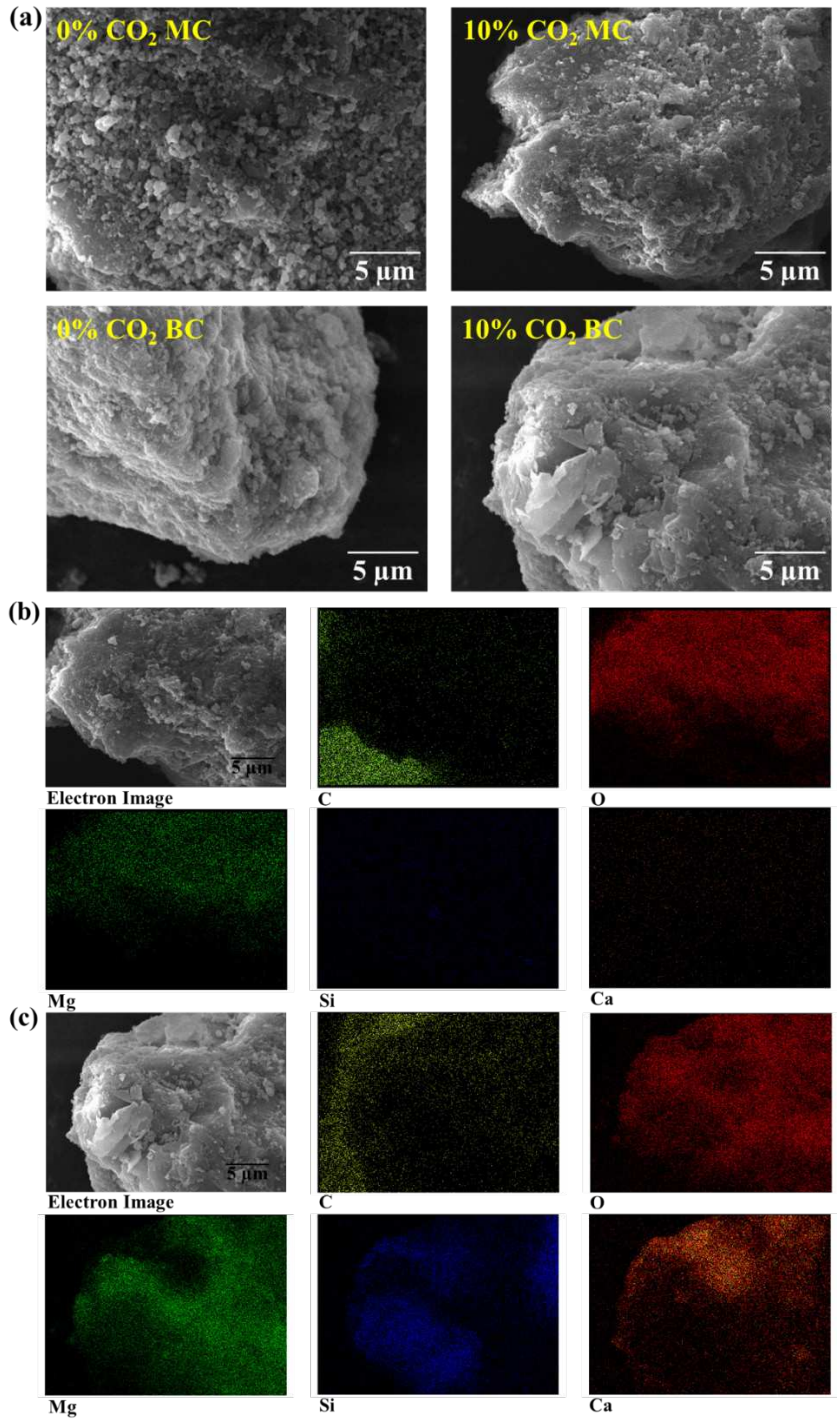
350 **Figure 9.** XRD diffractograms and Q-XRD analysis of MC pastes with different CO₂
 351 concentrations and curing time: (a) XRD diffractograms of 1-d cured pastes with different CO₂
 352 concentrations; (b) XRD diffractograms of pastes cured at 10% CO₂ concentration with
 353 different curing times; (c) Q-XRD analysis of 1-d cured pastes with different CO₂
 354 concentrations; (d) Q-XRD analysis of pastes cured at 10% CO₂ concentration with different
 355 curing times.

356

357 Figure 9a illustrates that the XRD peak due to Mg(OH)₂ at 18.7° 2θ gradually decreased with
 358 increasing CO₂ concentration, whereas the content of MgCO₃ increased. In samples exposed
 359 to 100% CO₂, the peaks of nesquehonite appeared at 23.4° and 34.5° 2θ. The Q-XRD results
 360 further suggested that a high content of amorphous phase (approximately 31.2%) existed in
 361 70% and 100% CO₂ cured samples, and HMCs were the dominant components of amorphous
 362 phase based on the TGA results. From Figure 9b&d, the content of MgCO₃ and amorphous
 363 HMCs in 10% CO₂ samples continually increased with curing time. The amorphous content

364 reached 30.2% after 7-d curing at 10% CO₂, which is further evidence that low-concentration
365 CO₂ and sufficient curing time can still promote carbonation of MC products.

366



367

368

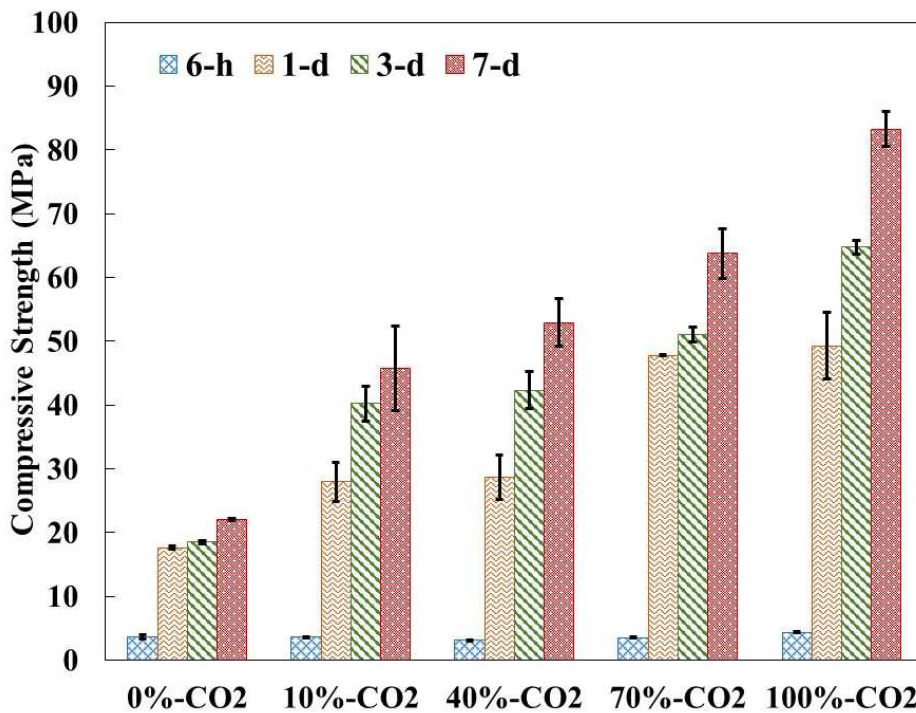
369 **Figure 10.** SEM images and elemental mapping of MC and BC samples: (a) SEM image of 1-
370 d 0% CO₂ and 10% CO₂ cured MC and BC samples, (b) element mapping of 1-d 10% CO₂
371 cured MC samples, (c) element mapping of 1-d 10% CO₂ cured BC samples.

372

373

374 As illustrated in SEM images (Figure 10a), 1-d air (0% CO₂) cured MC samples had a rough
375 surface with a large number of spherical particles, which may be unreacted MgO particles [36].
376 By comparison, 1-d 10% CO₂ cured MC samples showed a dense structure, suggesting
377 favourable MC hydration and carbonation. Elemental mapping (Figure 10b) showed that Mg
378 and O were the predominant elements and only a low content of C existed in 1-d 10% CO₂
379 cured samples. Therefore, under low-concentration CO₂ curing conditions, a long time CO₂
380 curing is required to achieve a high degree of carbonation.

381



382

383 **Figure 11.** Compressive strengths of MC pastes under CO₂ curing with various CO₂
384 concentrations.

385

386 Figure 11 shows that after 6-h CO₂ curing, all the MC blocks showed comparable compressive
387 strengths (3.74-4.41 MPa). After 1-d CO₂ curing, the differences in compressive strength were
388 significant among different CO₂ cured samples. The 1-d 10% CO₂ cured samples showed
389 higher strength than 7-d air cured samples, and strength increased with the increase of CO₂

390 concentration because of the favourable carbonation as indicated by the TGA and Q-XRD
391 results (Figure 8 and 9). It should be noted that extended 10% CO₂ curing (7-d) yielded similar
392 strengths to 1-d 100% CO₂ cured samples. These findings demonstrate that a low CO₂
393 concentration gas (e.g., industrial exhaust) could be used to accelerate carbonation of MC-
394 based materials, although it needs a relatively long curing time compared to 100% CO₂ curing.

395

396 **3.2.2 BC Pastes**

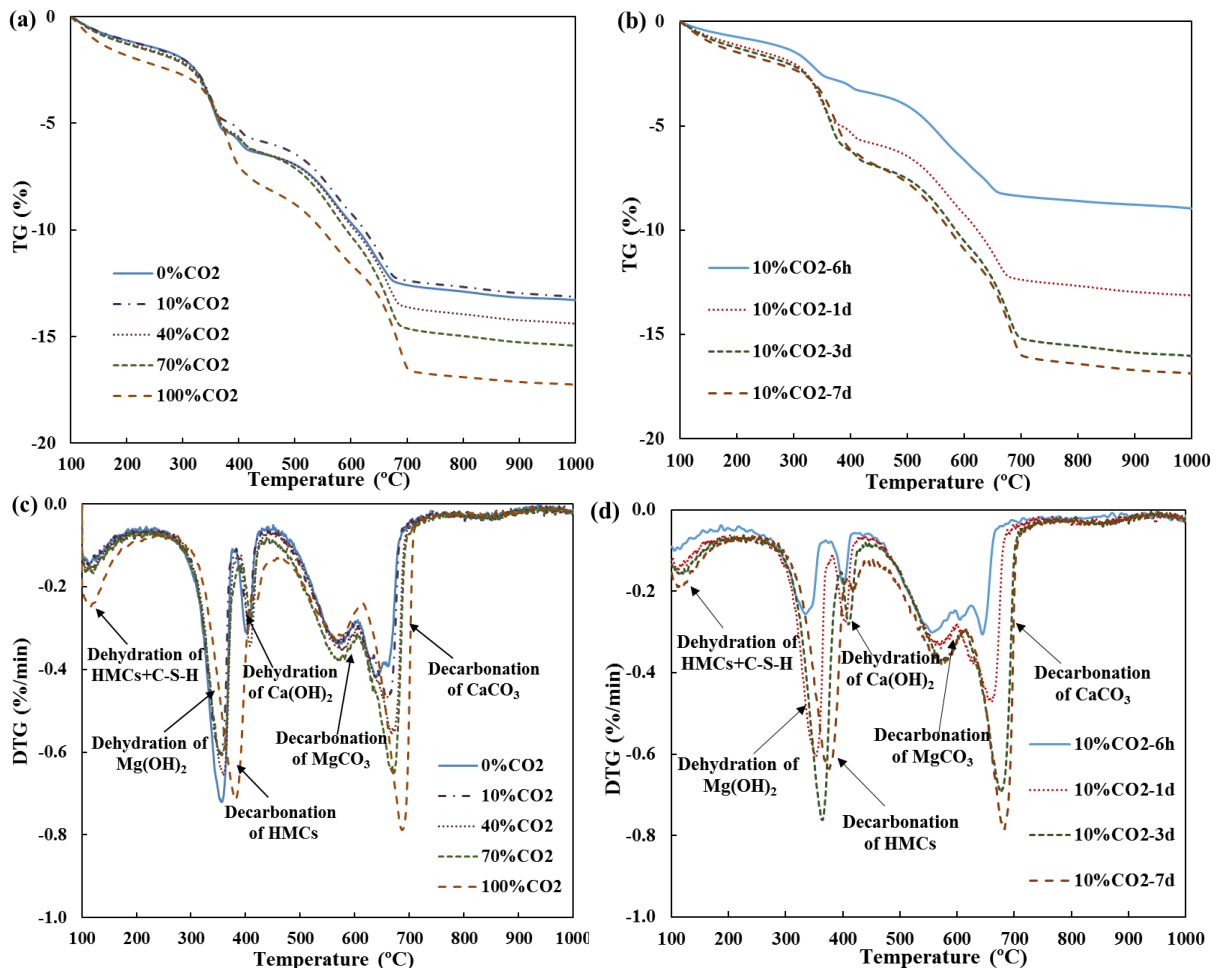
397 Figure 12a illustrates that the mass loss of 1-d carbonated BC samples increased with an
398 increase of CO₂ concentration, and that the Ca(OH)₂ in BC systems was preferentially
399 carbonated into CaCO₃. The obvious peaks of Ca(OH)₂ could be observed for the lower-CO₂
400 cured samples, whereas Ca(OH)₂ peaks nearly disappeared and Mg(OH)₂ was partially
401 transformed into HMCs and MgCO₃ in 100% CO₂ cured samples (Figure 12c). The contents
402 of carbonates increased with CO₂ curing time. The extent of calcite decomposition measured
403 by TG for 10%-CO₂ cured samples increased from 4.8 wt% (6-h curing) to 9.0 wt% (7-d
404 curing). The peak of Mg(OH)₂ (370 °C) was also shifted to a temperature indicative of HMCs
405 (390 °C) after 7-d CO₂ curing (Figure 12d). Thus, long-term curing at low concentration CO₂
406 curing could also accelerate the carbonation of BC.

407

408 As shown in Figure 13a&c, with the increase of CO₂ concentration the content of Ca(OH)₂ in
409 1-d cured BC samples gradually decreased, however, Mg(OH)₂ content was maintained at
410 approximately 8.6%, except in 100% CO₂ cured samples, suggesting that only Ca(OH)₂ was
411 partially carbonated in the early stages. From Figure 13b&d, the contents of Ca(OH)₂ and
412 Mg(OH)₂ in 10% CO₂ cured samples continually decreased with increasing curing time, while
413 the contents of calcite and amorphous phase accordingly increased. After 7-d curing at 10%
414 CO₂, the contents of calcite and amorphous phase (C-S-H, calcium aluminate hydrates,

415 amorphous HMCs, poorly crystalline carbonates, etc.) reached 13.5% and 36.9%, respectively,
 416 which were also comparable to the contents of these phases in 1-d 100% CO₂ cured BC
 417 samples.

418



419

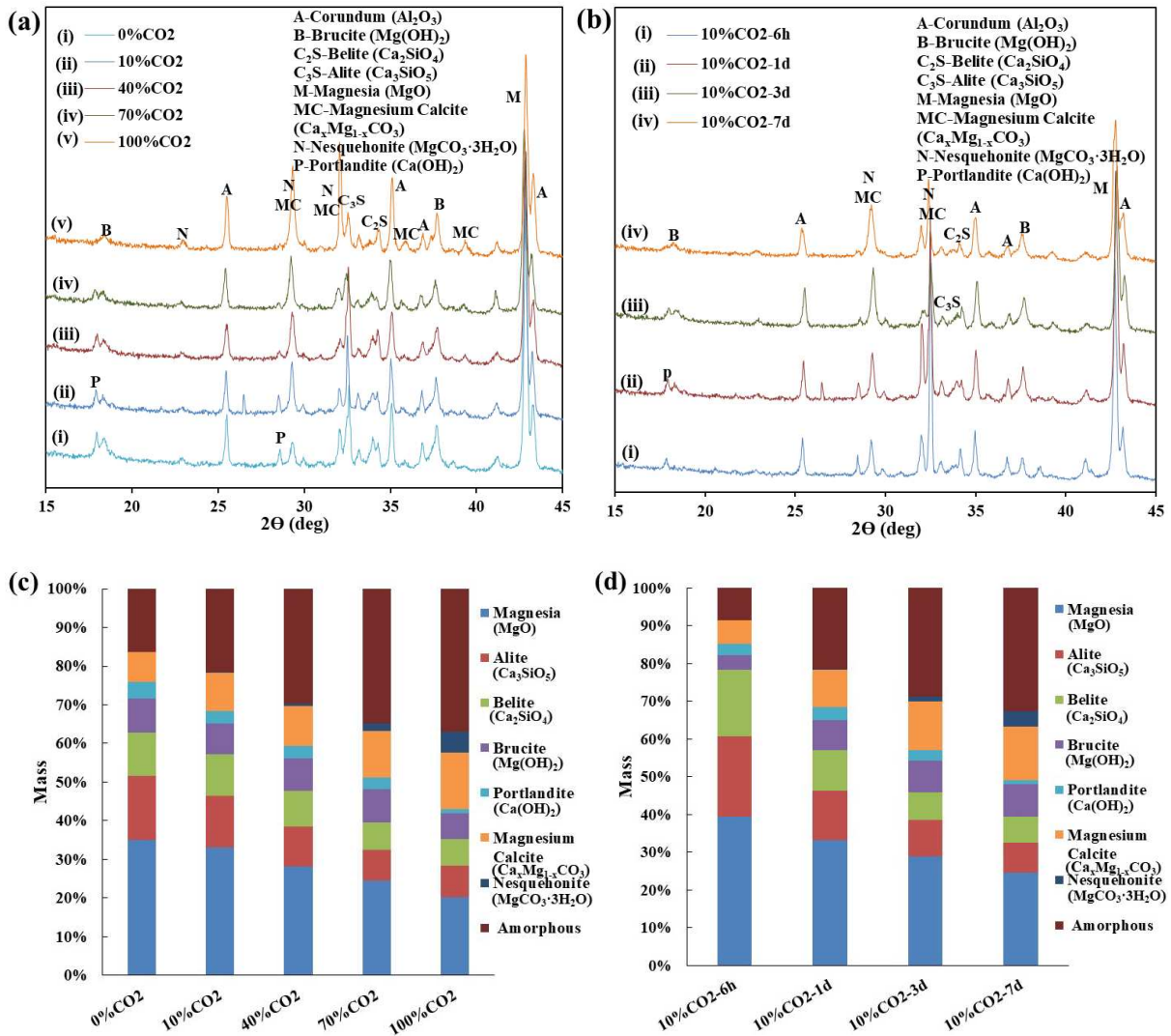
420

421 **Figure 12.** TGA of BC pastes: (a) TG curves of 1-d pastes as a function of CO₂ concentrations;
 422 (b) TG curves of pastes at cured at 10% CO₂ for different durations; (c) DTG curves
 423 corresponding to (a); (d) DTG curves corresponding to (b).

424

425 From SEM imaging (Figure 10a), 1-d air cured BC samples had a rough surface but the
 426 spherical particles on the surface were not remarkable. In contrast to the MC samples, most of
 427 MgO particles in the BC samples might be enclosed by PC hydrates (e.g., C-S-H gel) [37].
 428 After 1-d 10% CO₂ curing, a smooth and dense surface could be observed for BC samples.
 429 Elemental mapping (Figure 10c) illustrated that 1-d 10% CO₂ cured BC samples have a higher

430 content of carbon than the corresponding MC samples, and the carbon overlapped with the Ca-
 431 rich areas. This provides further evidence that CO₂ reacted with Ca²⁺ preferentially due to the
 432 relatively high solubility and reactivity of Ca(OH)₂. Therefore, the PC in the BC system played
 433 an important role in promoting the early-term carbonation in low-CO₂ concentration condition.
 434

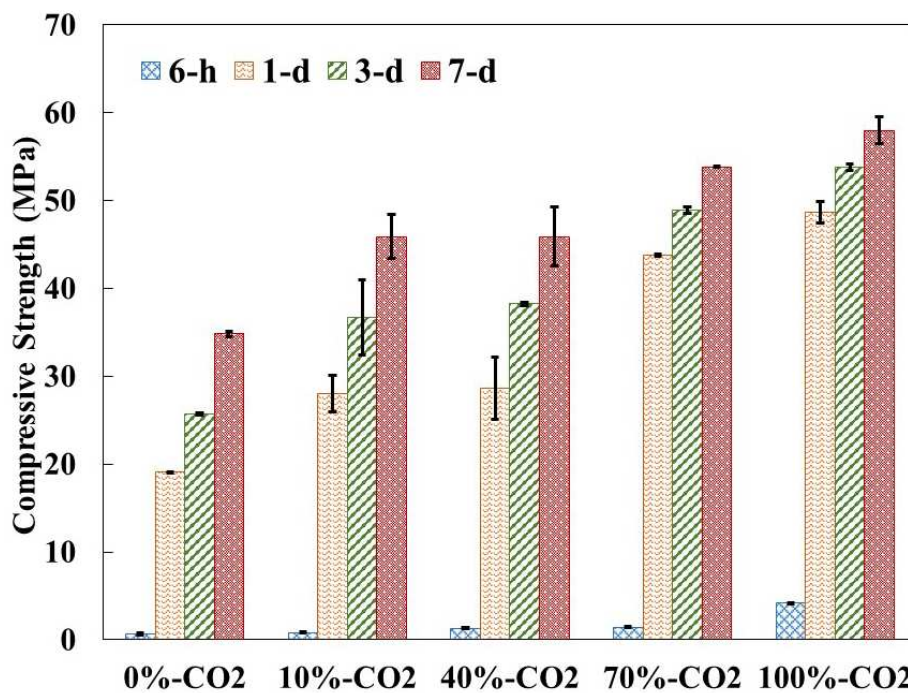


436
 437 **Figure 13.** XRD diffractograms and Q-XRD analysis of BC pastes with different CO₂
 438 concentrations and curing times: (a) XRD diffractograms of 1-d cured pastes with different
 439 CO₂ concentrations; (b) XRD diffractograms of pastes cured at 10% CO₂ concentration with
 440 different curing times; (c) Q-XRD analysis of 1-d cured pastes with different CO₂
 441 concentrations; (d) Q-XRD analysis of pastes cured at 10% CO₂ concentration with different
 442 curing times.

443
 444 As shown in Figure 14, BC samples with high-concentration CO₂ curing presented higher

445 compressive strengths at different curing durations due to their higher carbonation rate.
 446 However, the lower CO₂ concentration curing still effectively accelerated carbonation.
 447 Compared to 7-d air cured samples, the 7-d 10% CO₂ cured samples showed 31.8%
 448 enhancement of compressive strength, which gave comparable strength with 1-d 100% CO₂
 449 cured samples. Therefore, low CO₂ concentration gas under flow-through conditions is also
 450 effective for accelerating the carbonation of BC-based materials.

451



452

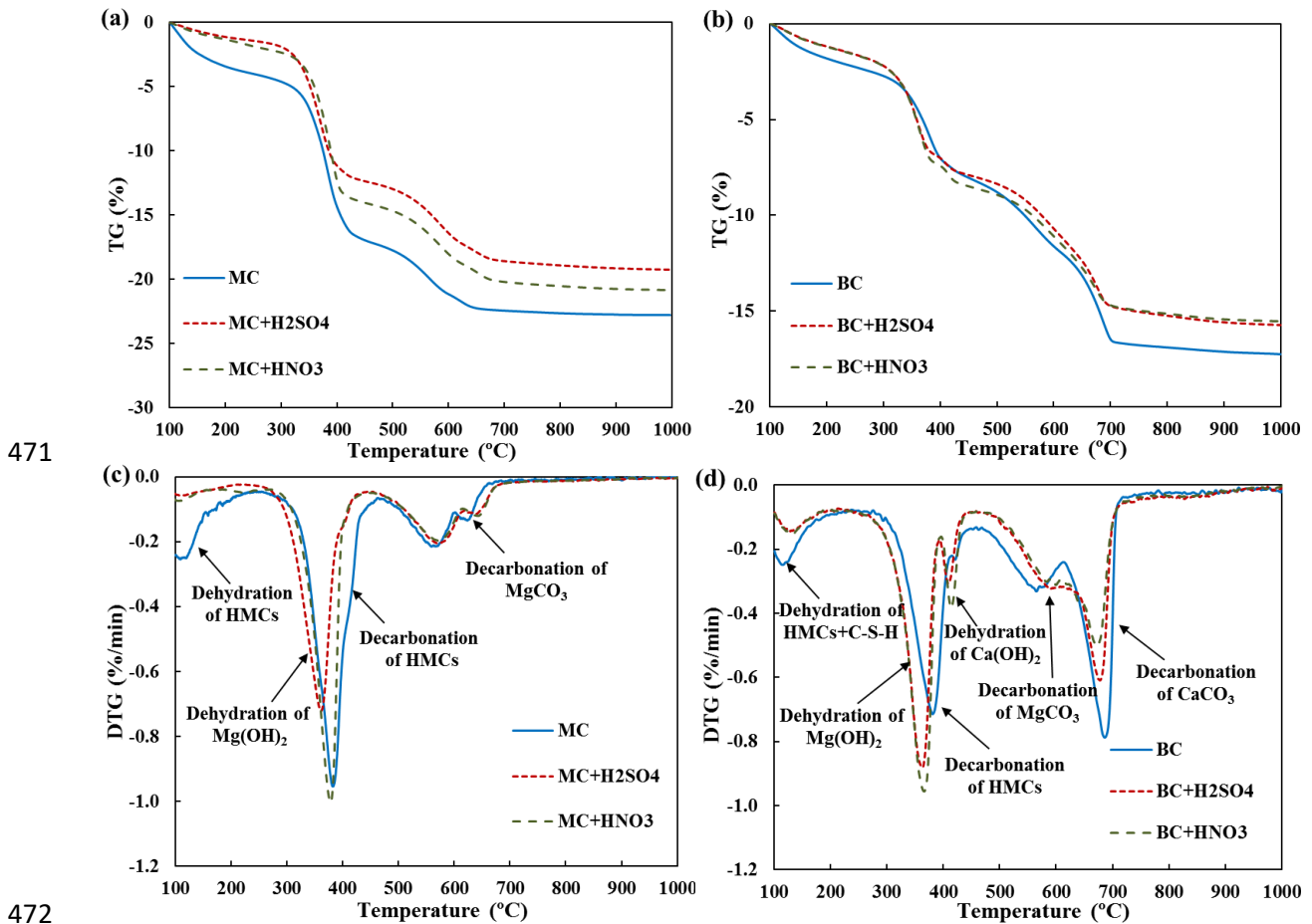
453 **Figure 14.** Compressive strength of BC pastes cured under CO₂ with various CO₂
 454 concentrations.

455

456 3.3 Effect of Acids on Accelerated Carbonation

457 The TG results (Figure 15) indicate that the mass loss of hydrates and carbonates of 1-d CO₂
 458 cured MC samples significantly reduced in the presence of H₂SO₄ and HNO₃ as representative
 459 acids that may be observed in flue gases. The carbonation of the MC samples was significantly
 460 inhibited by the acidic conditions. As the acidities of H₂SO₄ and HNO₃ are stronger than
 461 H₂CO₃, carbonates could not displace SO₄²⁻ and NO₃⁻ from the MC samples as the stronger
 462 acids reacted preferentially with the solid alkaline phases [38]. The sulphation and nitration are

463 irreversible processes [39]. It should be noted that the molar concentration of H_2SO_4 tested here
 464 was 5.83 times higher than the value of HNO_3 based on an investigation in power plant exhaust
 465 (in Supplementary Information). From Figure 15c, compared to acid-free MC samples, the
 466 mass loss in TG attributed to HMCs decomposition in acid-incorporated samples significantly
 467 reduced, especially for the H_2SO_4 -MC samples. Thus, in actual application, a high
 468 concentration of SO_2 in the exhaust may have a remarkable inhibitive effect on MC
 469 carbonation.
 470



473 **Figure 15.** TGA of 1-d 100% CO_2 cured pastes with H_2SO_4 and HNO_3 addition: (a) TG curves
 474 of MC pastes with addition of acids; (b) TG curves of BC pastes with addition of acids; (c)
 475 DTG curves corresponding to (a); (d) DTG curves corresponding to (b).

476

477

478 In the TG analysis of the 1-d CO₂ cured BC system (Figure 15b&d), the presence of H₂SO₄
479 and HNO₃ resulted in an 11% reduction in the mass loss of the hydrates and carbonates,
480 respectively, which was lower than the reduction in the MC system. This suggests that BC
481 samples had better compatibility with acids compared to MC samples. The high concentration
482 of H₂SO₄ and low concentration of HNO₃ had a similar inhibitory effect on carbonation in the
483 BC system. The H₂SO₄ may react with Ca(OH)₂ to form CaSO₄, which is one of the
484 components in PC (Table 1). Thus, PC in the BC system may to some degree be able to
485 compensate for the adverse effect of H₂SO₄.

486

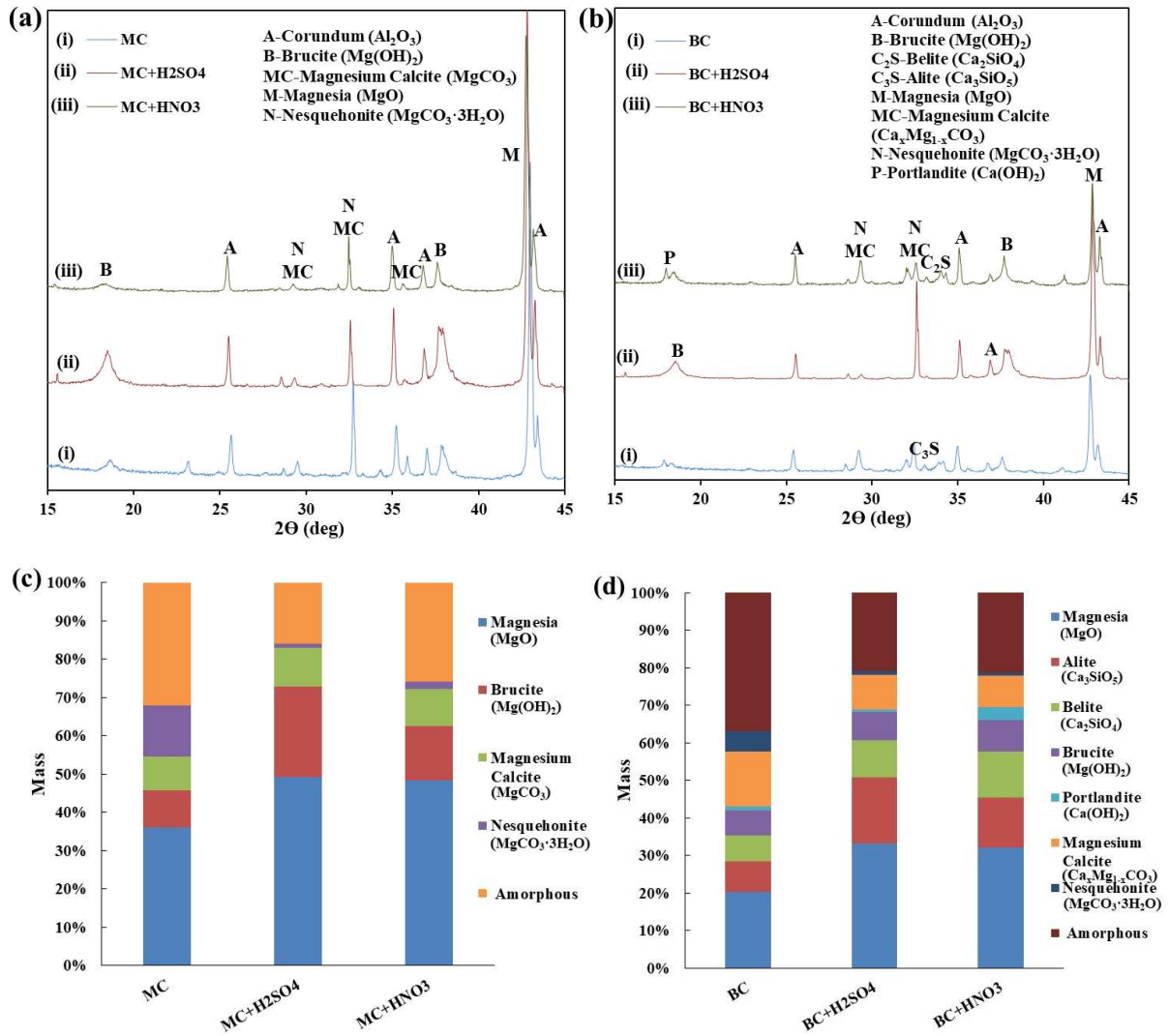
487 As shown in Figure 16, compared to acid-free CO₂ cured MC samples, the acid-incorporated
488 samples contained a higher content of uncarbonated Mg(OH)₂, particularly for the H₂SO₄
489 samples. Q-XRD analyses further indicated that the presence of H₂SO₄ and HNO₃ reduced the
490 content of nesquehonite by 90.3% and 84.7%, respectively, and also the amorphous content
491 decreased. In the BC system, the acids also resulted in the increase of unreacted raw materials
492 and decrease of carbonates and hydrates (crystalline or amorphous phases). In particular, the
493 presence of H₂SO₄ in BC samples consumed Ca(OH)₂ and decreased the content of calcite.

494

495 From Figure 17, the incorporation of acids decreased the compressive strength of 1-d air cured
496 MC samples by approximately 37.8% due to its inhibitory effect on the hydrates formation.
497 For CO₂ cured samples, the addition of H₂SO₄ and HNO₃ led to 64.0% and 44.5% strength
498 reduction, respectively. This suggests that acids significantly inhibited the carbonation of MC,
499 especially for high-concentration H₂SO₄, which was in accordance with TGA and XRD results.
500 Conversely, the incorporation of H₂SO₄ and HNO₃ showed a negligible effect on the strength
501 reduction in the air cured BC system, but reduced the strength of CO₂ cured BC samples by

502 40.8% and 39.8%, respectively. Hence, compared to the MC system, the BC system had a better
 503 compatibility with acids. To ensure effective accelerated carbonation, desulphurisation and
 504 acid gas neutralisation in the exhaust gas should be considered before employing this gas as a
 505 flow-through CO₂ curing source for MC and BC systems.

506

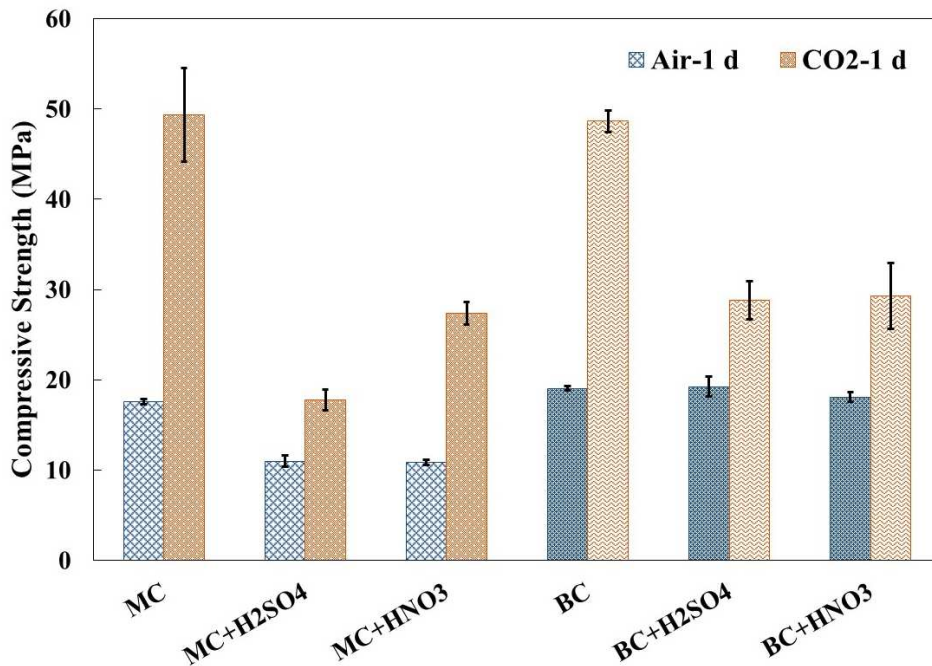


507

508

509 **Figure 16.** XRD diffractograms and Q-XRD analysis of 1-d 100% CO₂ cured pastes with
 510 H₂SO₄ and HNO₃ addition: (a) XRD diffractograms of MC pastes with addition of acids; (b)
 511 XRD diffractograms of BC pastes with addition of acids; (c) Q-XRD analysis of MC pastes
 512 with addition of acids; (d) Q-XRD analysis of BC pastes with addition of acids.

513



514 **Figure 17.** Compressive strengths of 1-d air and 100% CO₂ cured MC and BC samples with
 515 addition of acids.
 516
 517

518 4. Conclusions

519 This study investigated the accelerated carbonation of low-carbon reactive MgO cement and
 520 ordinary Portland cement blends under flow-through CO₂ gas conditions. The experimental
 521 results showed that flow-through CO₂ curing at a relatively low concentration (simulated
 522 industrial exhaust) effectively accelerated carbonation and enhanced compressive strength of
 523 single MgO cement (MC) pastes and binary cement (BC) pastes. High relative humidity (98%)
 524 was favourable for CO₂ curing of the MC system due to the high water consumption of MC for
 525 the generation and dissolution of Mg(OH)₂, whereas low humidity (50%) was an optimal
 526 parameter for BC paste carbonation. The addition of Portland cement (PC) into the BC system
 527 boosted early-term carbonation rate due to the rapid formation and dissolution of Ca(OH)₂.
 528 Despite the lower carbonation rate at 10% CO₂ concentration, the carbonation degree and
 529 compressive strength of 7-d 10% CO₂ cured samples were comparable to the values obtained
 530 for 1-d 100% CO₂ cured samples. Nevertheless, the presence of acids replicating the acid gases
 531 in industrial exhausts had inhibitory effects on the carbonation and hydration of MC pastes. By

532 comparison, the PC included in the BC system have a better compatibility with SO_4^{2-} and partly
533 relieved the inhibitory effect. In future studies, the effect of NO_2 and SO_2 from the exhaust gas
534 on the carbonation and hydration of cement-based materials should be further evaluated, and
535 the flow-through curing system should be optimised to enhance the recovery rate of CO_2 from
536 the exhaust gas. Overall, this study demonstrated that low-carbon MgO-based cement can be a
537 promising material for sequestering and utilising CO_2 from industrial exhausts to promote the
538 development of negative-emission technologies.

539

540 **Acknowledgement**

541 The authors appreciate the financial support from the Hong Kong Research Grants Council
542 (PolyU 15223517 and E-PolyU503/17) for this study.

543

544 **References**

545 [1] K.L. Scrivener, V.M. John, E.M. Gartner, Eco-efficient cements: Potential economically
546 viable solutions for a low- CO_2 cement-based materials industry, *Cem. Concr. Res.* 114 (2018)
547 2–26.

548 [2] N.T. Dung, C. Unluer, Development of MgO concrete with enhanced hydration and
549 carbonation mechanisms, *Cem. Concr. Res.* 103 (2018) 160–169.

550 [3] Y. Geng, Z. Wang, L. Shen, J. Zhao, Calculating of CO_2 emission factors for Chinese
551 cement production based on inorganic carbon and organic carbon, *J. Clean. Prod.* 217 (2019)
552 503–509.

553 [4] L. Wang, K. Yu, J.S. Li, D.C.W. Tsang, C.S. Poon, J.C. Yoo, K. Baek, S. Ding, D. Hou,
554 J.G. Dai, Low-carbon and low-alkalinity stabilization/solidification of high-Pb contaminated
555 soil, *Chem. Eng. J.* 351 (2018) 418–427.

- 556 [5] A. Al-Tabbaa, Reactive magnesia cement, in: *Eco-Efficient Concrete*. Woodhead
557 Publishing (2013) 523–543.
- 558 [6] T. Zhang, X. Liang, C. Li, M. Lorin, Y. Li, L.J. Vandeperre, C.R. Cheeseman, Control of
559 drying shrinkage in magnesium silicate hydrate (M-S-H) gel mortars, *Cem. Concr. Res.* 88
560 (2016) 36–42.
- 561 [7] H. Dong, E.H. Yang, C. Unluer, F. Jin, A. Al-Tabbaa, Investigation of the properties of
562 MgO recovered from reject brine obtained from desalination plants, *J. Clean. Prod.* 196 (2018)
563 100–108.
- 564 [8] S.A. Walling, J.L. Provis, Magnesia-Based Cements: A Journey of 150 Years, and Cements
565 for the Future?, *Chem. Rev.* 116 (2016) 4170–4204.
- 566 [9] Y. Tan, H. Yu, Y. Li, W. Bi, X. Yao, The effect of slag on the properties of magnesium
567 potassium phosphate cement, *Constr. Build. Mater.* 126 (2016) 313–320.
- 568 [10] L. Wang, L. Chen, D.W. Cho, D.C.W. Tsang, J. Yang, D. Hou, K. Baek, H.W. Kua, C.S.
569 Poon, Novel synergy of Si-rich minerals and reactive MgO for stabilisation/solidification of
570 contaminated sediment, *J. Hazard. Mater.* 365 (2019) 695–706.
- 571 [11] L. Wang, L. Chen, D.C.W. Tsang, H.W. Kua, J. Yang, Y.S. Ok, S.M. Ding, D.Y. Hou,
572 C.S. Poon, The roles of biochar as green admixture for sediment-based construction products.
573 *Cem. Concr. Compos.* 104 (2019) 103348.
- 574 [12] C. Kuenzel, F. Zhang, V. Ferrándiz-Mas, C.R. Cheeseman, E.M. Gartner, The mechanism
575 of hydration of MgO-hydromagnesite blends, *Cem. Concr. Res.* 103 (2018) 123–129.
- 576 [13] L. Wang, I.K.M. Yu, D.C.W. Tsang, K. Yu, S. Li, C. Sun Poon, J.G. Dai, Upcycling wood
577 waste into fibre-reinforced magnesium phosphate cement particleboards, *Constr. Build. Mater.*
578 159 (2018) 54–63.

- 579 [14] L. Mo, F. Zhang, M. Deng, F. Jin, A. Al-Tabbaa, A. Wang, Accelerated carbonation and
580 performance of concrete made with steel slag as binding materials and aggregates, *Cem. Concr.*
581 *Compos.* 83 (2017) 138–145.
- 582 [15] F. Cao, M. Miao, P. Yan, Hydration characteristics and expansive mechanism of MgO
583 expansive agents, *Constr. Build. Mater.* 183 (2018) 234–242.
- 584 [16] B.J. Zhan, D.X. Xuan, C.S. Poon, C.J. Shi, Effect of curing parameters on CO₂ curing of
585 concrete blocks containing recycled aggregates, *Cem. Concr. Compos.* 71 (2016) 122–130.
- 586 [17] C. Shi, Z. Wu, Z. Cao, T.C. Ling, J. Zheng, Performance of mortar prepared with recycled
587 concrete aggregate enhanced by CO₂ and pozzolan slurry, *Cem. Concr. Compos.* 86 (2018)
588 130–138.
- 589 [18] N.T. Dung, C. Unluer, Carbonated MgO concrete with improved performance: The
590 influence of temperature and hydration agent on hydration, carbonation and strength gain,
591 *Cem. Concr. Compos.* 82 (2017) 152–164.
- 592 [19] S.A. Walling, J.L. Provis, A discussion of the papers “Impact of hydrated magnesium
593 carbonate additives on the carbonation of reactive MgO cements” and “Enhancing the
594 carbonation of MgO cement porous blocks through improved curing conditions”, by C. Unluer
595 & A. Al-Tabbaa, *Cem. Concr. Res.* 79 (2016) 424–426.
- 596 [20] M. Fernández Bertos, S.J.R. Simons, C.D. Hills, P.J. Carey, A review of accelerated
597 carbonation technology in the treatment of cement-based materials and sequestration of CO₂.,
598 *J. Hazard. Mater.* 112 (2004) 193–205.
- 599 [22] L. Mo, F. Zhang, D.K. Panesar, M. Deng, Development of low-carbon cementitious
600 materials via carbonating Portland cement–fly ash–magnesia blends under various curing
601 scenarios: a comparative study, *J. Clean. Prod.* 163 (2017) 252–261.
- 602 [21] L. Pu, C. Unluer, Durability of carbonated MgO concrete containing fly ash and ground
603 granulated blast-furnace slag, *Constr. Build. Mater.* 192 (2018) 403–415.

- 604 [23] L. Mo, D.K. Panesar, Effects of accelerated carbonation on the microstructure of Portland
605 cement pastes containing reactive MgO, *Cem. Concr. Res.* 42 (2012) 769–777.
- 606 [24] S. Ruan, C. Unluer, Influence of mix design on the carbonation, mechanical properties
607 and microstructure of reactive MgO cement-based concrete, *Cem. Concr. Compos.* 80 (2017)
608 104–114.
- 609 [25] R. Zhang, D.K. Panesar, Mechanical properties and rapid chloride permeability of
610 carbonated concrete containing reactive MgO, *Constr. Build. Mater.* 172 (2018) 77–85.
- 611 [26] L. Wang, S.S. Chen, D.C.W. Tsang, C.S. Poon, J.G. Dai, CO₂ curing and fibre
612 reinforcement for green recycling of contaminated wood into high-performance cement-
613 bonded particleboards, *J. CO₂ Util.* 18 (2017) 107–116.
- 614 [27] R. Zhang, N. Bassim, D.K. Panesar, Characterization of Mg components in reactive MgO
615 - Portland cement blends during hydration and carbonation, *J. CO₂ Util.* 27 (2018) 518–527.
- 616 [28] W. Shen, L. Cao, Q. Li, Z. Wen, J. Wang, Y. Liu, R. Dong, Y. Tan, R. Chen, Is magnesia
617 cement low carbon? Life cycle carbon footprint comparing with Portland cement, *J. Clean.*
618 *Prod.* 131 (2016) 20–27.
- 619 [29] L. Wang, L. Chen, D.C.W. Tsang, J.S. Li, T.L.Y. Yeung, S. Ding, C.S. Poon, Green
620 remediation of contaminated sediment by stabilization/solidification with industrial by-
621 products and CO₂ utilization, *Sci. Total Environ.* 631–632 (2018) 1321–1327.
- 622 [30] I. Taniguchi, T. Yamada, Low energy CO₂ capture by electrodialysis, in: *Energy Proced.*
623 (2017) 1615–1620.
- 624 [31] D. Xuan, B. Zhan, C.S. Poon, A maturity approach to estimate compressive strength
625 development of CO₂-cured concrete blocks, *Cem. Concr. Compos.* 85 (2018) 153–160.
- 626 [32] BS EN 12390, Testing Hardened Concrete Compressive Strength of Test Specimens.
627 British Standards Institution, London, UK (2009).

628 [33] L. Han, X. Li, J. Bai, F. Xue, Y. Zheng, C. Chu, Effects of flow velocity and different
629 corrosion media on the in vitro bio-corrosion behaviours of AZ31 magnesium alloy, *Mater.*
630 *Chem. Phys.* 217 (2018) 300–307.

631 [34] L. Wang, S.S. Chen, D.C.W. Tsang, C.S. Poon, K. Shih, Recycling contaminated wood
632 into eco-friendly particleboard using green cement and carbon dioxide curing, *J. Clean. Prod.*
633 137 (2016) 861–870.

634 [35] C. Unluer, A. Al-Tabbaa, Enhancing the carbonation of MgO cement porous blocks
635 through improved curing conditions, *Cem. Concr. Res.* 59 (2014) 55–65.

636 [36] L. Wang, L. Chen, D.W. Cho, D.C.W. Tsang, J. Yang, D. Hou, K. Baek, H.W. Kua, C.S.
637 Poon, Novel synergy of Si-rich minerals and reactive MgO for stabilisation/solidification of
638 contaminated sediment, *J. Hazard. Mater.* (2019) 695–706.

639 [37] D.K. Panesar, L. Mo, Properties of binary and ternary reactive MgO mortar blends
640 subjected to CO₂ curing, *Cem. Concr. Compos.* 38 (2013) 40–49.

641 [38] H. Chen, C. Zhao, Q. Ren, Feasibility of CO₂/SO₂ uptake enhancement of calcined
642 limestone modified with rice husk ash during pressurized carbonation, *J. Environ. Manage.* 93
643 (2012) 235–244.

644 [39] A. Coppola, A. Esposito, F. Montagnaro, G. De Tommaso, F. Scala, P. Salatino, Effect of
645 exposure to SO₂ and H₂O during the carbonation stage of fluidised bed calcium looping on the
646 performance of sorbents of different nature, *Chem. Eng. J.* (2018).

647

648

649 **Supplementary Information**

650 **Calculation of H₂SO₄ and HNO₃ dosage in MC and BC pastes**

651 Based on our field investigation, the typical CO₂, SO₂, and NO₂ concentrations in thermal
652 power plant exhaust are 7-12%, 3500-4500 mg Nm⁻³, and 400-600 mg Nm⁻³, respectively. In
653 this study, we assumed that CO₂, SO₂, and NO₂ concentrations from the thermal power plant
654 are **10%**, 4000 mg Nm⁻³, and 500 mg Nm⁻³, respectively; these **SO₂ and NO₂** concentrations
655 are equivalent to **0.14% and 0.024%** in the exhaust, respectively.

656

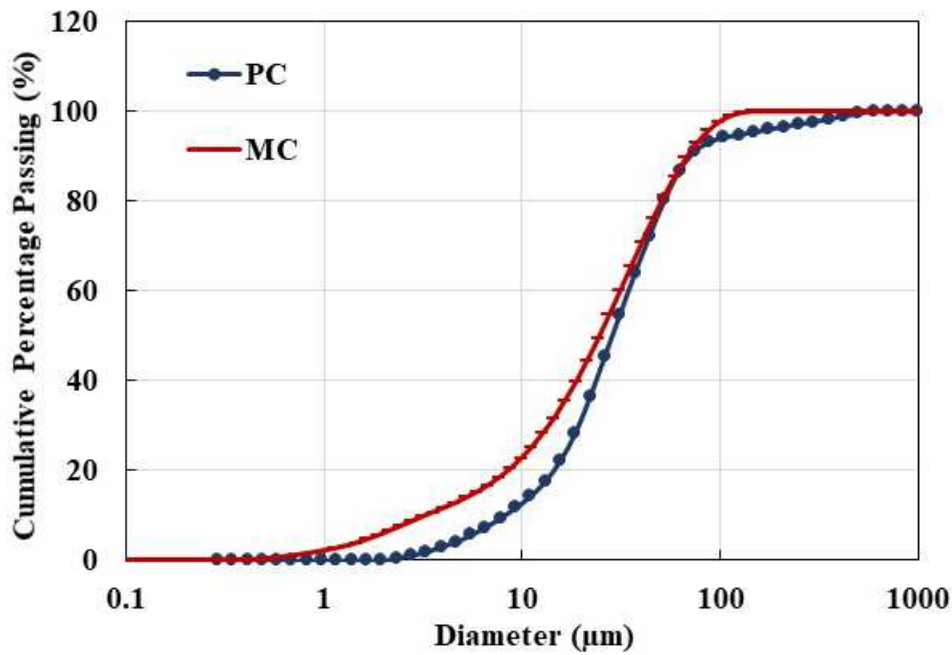
657 According to TGA results, the **CO₂ absorption extent is 4%** in the MC paste (CO₂
658 concentration of 100%, humidity of 98%). The reaction between acid gas and alkaline paste is
659 a diffusion-controlled reaction. Here, we assume that diffusion rates of acid gas are the same
660 as the rate of diffusion of CO₂ gas. Thus, the absorption extent of acid gas depends on the
661 concentration of acid gas in the exhaust.

662 We can calculate the dosage of H₂SO₄ and HNO₃ as follows:

663
$$m_{H_2SO_4} = 4\% \frac{0.14}{10} \times \frac{98}{44} \times 100\% = 0.125\% g$$

664
$$m_{HNO_3} = 4\% \frac{0.024}{10} \times \frac{63}{44} \times 100\% = 0.0137\% g$$

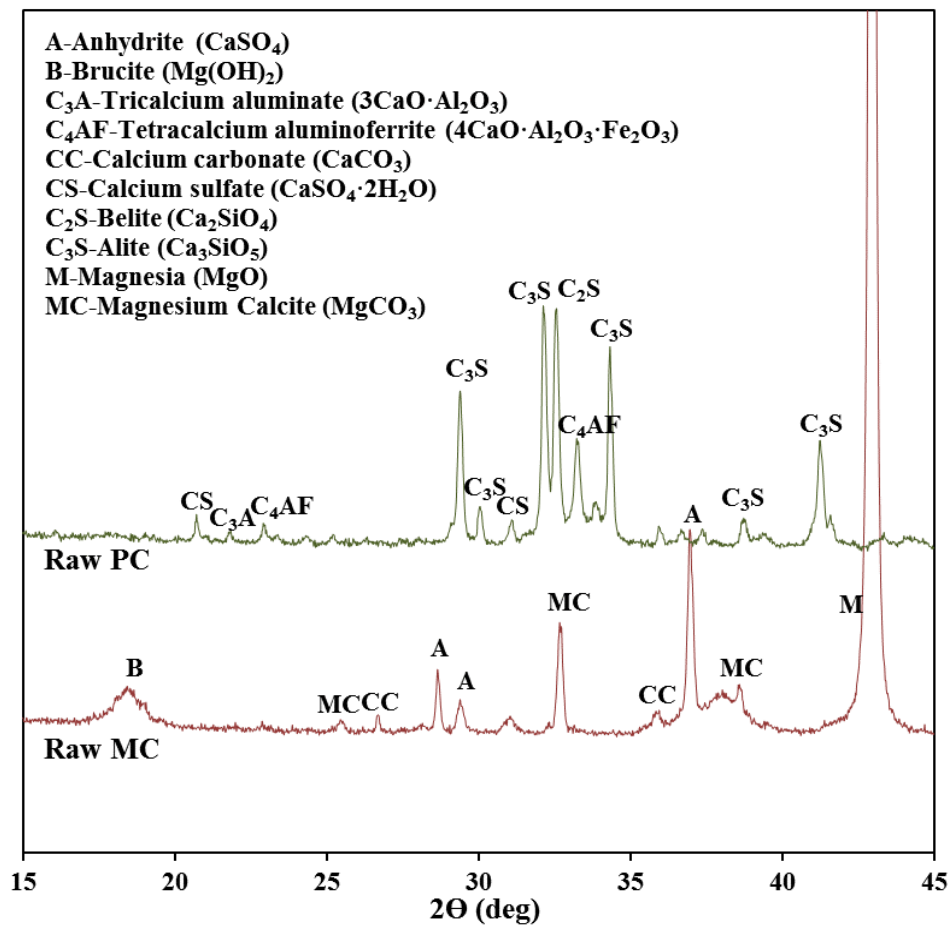
665 Therefore, **H₂SO₄ (0.125 wt% of paste)** and **HNO₃ (0.0137 wt% of paste)** are added in the
666 specified mixtures, respectively.



667

668 **Figure S1.** Particle size distribution of raw PC and MC.

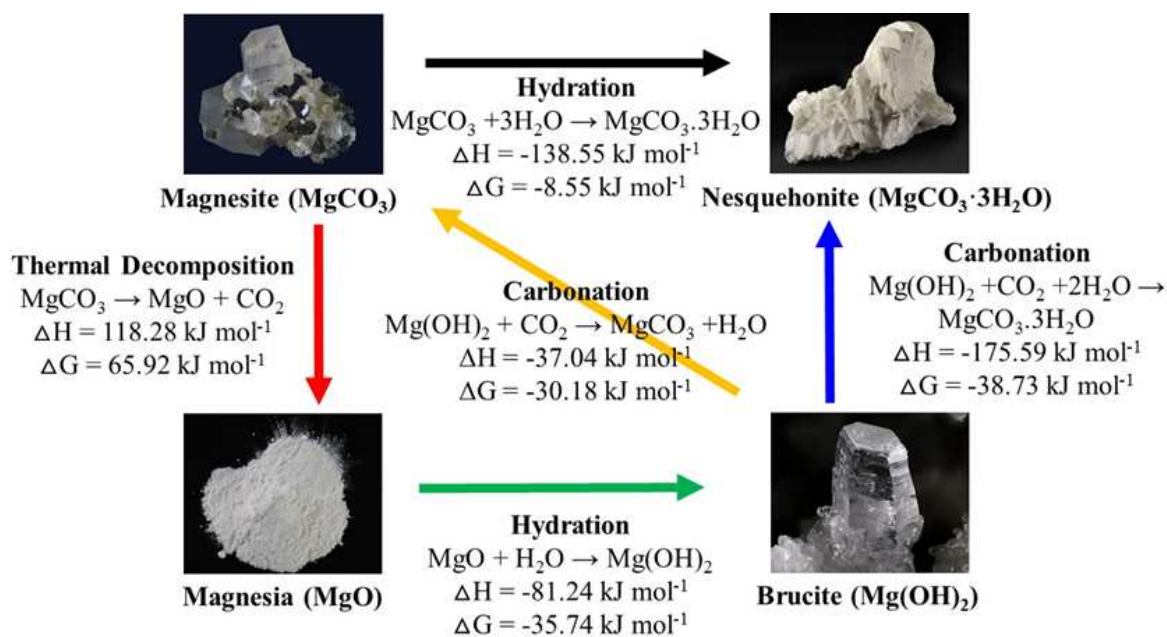
669



670

671 **Figure S2.** XRD diffractograms of raw PC and MC.

672



673

674 **Figure S3.** Thermodynamic reactions of the family of magnesium oxides/carbonates (Wang et
 675 al., 2016). Copyright Elsevier.

676

677 **Reference:**

678 Wang, L., Chen, S.S., Tsang, D.C.W., Poon, C.S., Shih K.M., 2016. Recycling contaminated
 679 wood into eco-friendly particleboard using green cement and carbon dioxide curing. J.
 680 Clean. Prod. 137, 861-870.

**ELSEVIER LICENSE
TERMS AND CONDITIONS**

Sep 19, 2019

This Agreement between Mr. liang chen ("You") and Elsevier ("Elsevier") consists of your license details and the terms and conditions provided by Elsevier and Copyright Clearance Center.

| | |
|---|--|
| License Number | 4672520469961 |
| License date | Sep 19, 2019 |
| Licensed Content Publisher | Elsevier |
| Licensed Content Publication | Journal of Cleaner Production |
| Licensed Content Title | Recycling contaminated wood into eco-friendly particleboard using green cement and carbon dioxide curing |
| Licensed Content Author | Lei Wang,Season S. Chen,Daniel C.W. Tsang,Chi-Sun Poon,Kaimin Shih |
| Licensed Content Date | Nov 20, 2016 |
| Licensed Content Volume | 137 |
| Licensed Content Issue | n/a |
| Licensed Content Pages | 10 |
| Start Page | 861 |
| End Page | 870 |
| Type of Use | reuse in a journal/magazine |
| Requestor type | academic/educational institute |
| Intended publisher of new work | Elsevier |
| Portion | figures/tables/illustrations |
| Number of figures/tables/illustrations | 1 |
| Format | both print and electronic |
| Are you the author of this Elsevier article? | Yes |
| Will you be translating? | No |
| Original figure numbers | figures 1 |
| Title of the article | Accelerated Carbonation of Reactive MgO and Portland Cement Blends Under Flowing CO ₂ Gas |
| Publication new article is in | Cement and Concrete Composites |
| Publisher of the new article | Elsevier |
| Author of new article | Lei Wang, Liang Chen, John L. Provis, Daniel C.W. Tsang, Chi-Sun Poon |
| Expected publication date | Jul 2019 |
| Estimated size of new article (number of pages) | 21 |
| Requestor Location | Mr. liang chen ZS922, 9/F, Block Z, polyu |

<https://s100.copyright.com/AppDispatchServlet>

1/6