| 1 | Potassium based sorbents from fly ash for high temperature CO ₂ capture |
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| 9 | |
| 10 | Abstract |
| 11 12 13 14 15 16 17 18 19 20 21 22 23 24 | Potassium-fly ash (K-FA) sorbents were investigated for high temperature CO ₂ sorption. K-FAs were synthetised using coal fly ash as source of silica and aluminium. The synthetised materials were also mixed with Li ₂ CO ₃ and Ca(OH) ₂ to evaluate their effect on CO ₂ capture. Temperature strongly affected the performance of the K-FA sorbents, resulting in a CO ₂ uptake of 1.45 mmol CO ₂ /g sorbent for K-FA 1:1 at 700°C. The CO ₂ sorption was enhanced by the presence of Li ₂ CO ₃ (10 wt%), with the K-FA 1:1 capturing 2.38 mmol CO ₂ /g sorbent at 700°C in 5 min. This sorption was found to be similar to previously developed Li-Na-FA (2.54 mmol/g) and Li-FA (2.4 mmol/g) sorbents. The presence of 10% Li ₂ CO ₃ also accelerated sorption and desorption. The results suggest that the increased uptake of CO ₂ and faster reaction rates in presence of K-FA can be ascribed to the formation of K-Li eutectic phase, which favours the diffusion of potassium and CO ₂ in the material matrix. The cyclic experiments showed that the K-FA materials maintained stable CO ₂ uptake and reaction rates over 10 cycles. |
| 25 | Keywords: Absorption, CO ₂ sorbents, Fly ash, Potassium, CCS |
| 26 | |
| 27 | Introduction |
| 28 | The requirement for mitigating climate change is supporting the development of CO ₂ capture |
| 29 | technologies. At the state of the art, CO ₂ absorption by liquid organic amines is the most |
| 30 | advanced technology (Leung at al., 2014). However, amine degradation in both absorption |
| 31 | and stripping steps and the potential emission of harmful compounds need to be addressed for |
| 32 | a widespread commercialisation of this technology (Manzoor et al., 2014; Vega et al., 2015). |
| 33 | The incorporation of amines into porous support has been proposed as alternative approach, |
| 34 | but the low amine utilization ratio and the loss of amine compounds in the regeneration step |
| 35 | required the development of new sorbents (Zhao et al., 2014). |
| 36 | A suitable sorbent for CO ₂ capture from flue gas should satisfy several important criteria to |

A suitable sorbent for CO_2 capture from flue gas should satisfy several important criteria to compete with the present technologies, including high sorption capacity, adequate sorption/desorption kinetics and stability, mechanical strength etc. (Sabouni et al., 2014). A large number of physical adsorbent materials have been considered for low temperature (<
150°C) CO₂ capture such as activated carbon, zeolites and metal organic frameworks (MOFs)
(Samanta et al., 2012).

 CO_2 capture at high temperatures (> 400°C) using solid sorbents with high selectivity and 42 regenerability has also been proposed as an alternative to low-temperature CO₂ capture due to 43 reduced efficiency penalties (Olivares-Marín et al., 2010). Due to their high decomposition 44 temperature (T > 800°C), alkaline ceramics, mainly lithium containing compounds have been 45 tested as CO₂ sorbents at high temperature (Rodríguez and Pfeiffer, 2008; Olivares-Marín et 46 al., 2010). The CO₂ sorption on ceramics starts with the carbonation of the sorbents surface 47 48 and continues with the diffusion of CO₂ through the carbonate external layer to react with the alkaline core (Rodríguez and Pfeiffer, 2008). So far, potassium has been mainly investigated 49 as CO₂ sorption promoter associated to other ceramics such as Li- or Na- based sorbents, 50 zeolites and alumina based CO₂ sorbents (Seggiani et al., 2013; Zhao et al., 2014; Olivares-51 Marín et al., 2010; Sanna et al., 2014; Sanna et al., 2015). 52

Amongst the wide range of materials tested, lithium silicate (Li₄SiO₄) and aluminate 53 (Li₅AlO₄) have shown the largest CO₂ sorption capacity and the fastest CO₂ sorption rate 54 over a wide range of temperatures and CO₂ concentrations (Olivares-Marín et al., 2010; Kato 55 et al., 2005, Flores-Martínez and Pfeiffer, 2015). It has been reported that the presence of 40 56 mol% K₂CO₃ at 600°C affects the CO₂ sorption capacity for Li-based sorbents (Li-FA) 57 58 prepared from fly ashes (FA), with a sorption capacity of 2.43 mmol CO₂/g sorbent (Olivares-Marín et al., 2010). Recently, sodium-fly ash (Na-FA) sorbents have also been 59 proposed as high temperature CO₂ sorbents. Maximum CO₂ uptake of Na-FA sorbent was 60 found to be 2 mmol CO₂/g sorbent in presence of 20% Li₂CO₃ additive (Sanna et al., 2014; 61 Sanna at al., 2015). Even if lithium based materials present very good CO₂ uptake 62 performance, it should be noted that lithium is considered a rare element in comparison to 63 sodium and potassium. It has been assessed that demand for lithium is growing at a fast rate 64 and that this rate exceeds projected availability by 25% (The Financialist, 2014). Therefore, 65 K-based sorbents can be considered not rare compared to lithium based materials. 66

The use of potassium carbonate, in addition to other alkali-metal materials was firstly studied by NASA for space applications (Onischak et al, 1978). Also, potassium has been previously identified as a good candidate for low temperature CO_2 sorption while utilizing an activated carbon as support (Hayashi et al., 1998). In addition, silver carbonate in combination with alkali metal silicate (including potassium), alkali metal carbonate (potassium and/or sodium)
have been evaluated as CO₂ sorbent (Nalette et al., 1992).

Thermodynamic analysis including enthalpy and free energy changes have been calculated 73 for both K-silicate sorption and regeneration reactions (Hoffman and Pennline, 2001). Results 74 for alkali-based sorbents are generally favourable in that the forward (CO₂ absorption) 75 reaction rate is typically much larger than the reverse reaction. Potassium carbonate was 76 found to be suitable for CO₂ capture at low absorption temperatures (less than 145°C), while 77 calcium oxide is more suitable for high absorption temperatures (less than 860°C) (Hoffman 78 79 and Pennline, 2001). For this reason, potassium-based sorbents have been proposed as effective ambient temperature CO₂ sorbent in confined spaces (Zhao et al., 2014). The CO₂ 80 sorption capacities were calculated as 0.87, 1.18, 0.34, 0.53, and 0.15 mmol CO₂/g for 81 K₂CO₃/Activate K_2CO_3/Al_2O_3 , $K_2CO_3/zeolite-5A$ 82 carbon, and $K_2CO_3/13X$ -zeolite, respectively. 83

Despite the fact that potassium carbonate has been widely tested as CO₂ capture promoter 84 (Olivares-Marín et al., 2010; Seggiani et al, 2011; Flores-Martínez and Pfeiffer, 2015; Sanna 85 et al, 2015), there is a lack of works on the potential use of K-silicates derived from fly ash as 86 high temperature CO₂ sorbents and their behaviour when used in presence of CO₂ sorption 87 88 promoters. Moreover, previous studies have only examined the CO₂ capture behaviours of alkali metal-based sorbents under conditions of 50–100°C, 5–20% CO₂ and 5–20% H_2O 89 (Zhao et al, 2011; Zhao et al, 2013; Zhao et al, 2014). It is unclear whether CO₂ sorption 90 capacity changes at higher temperatures. The aim of the present work was to produce pure 91 92 and Li-doped aluminosilicates from fly ash and assess their CO₂ sorption capacity under 93 different conditions.

94

95 **Experimental**

96 Fly ashes (FA) have been used as SiO_2 source and collected from a cyclone filter. The parent 97 sample was characterised by particles size distribution and XRF in previous work (A. Sanna 98 et al, 2015). The potassium based CO₂ sorbents were synthetised by mixing the FA with 99 K₂CO₃ (Acros Organics) at different K₂CO₃:SiO₂ (from FA) molar ratios (2:1 and 1:1) by 100 using an agate mortar and pestle. The mixed powders were calcined in a muffle furnace at 101 800°C for 8 hours. The sorbent synthesis is described as follow:

102 $K_2CO_3 + SiO_2 \text{ (from fly ash)} \rightarrow K_2SiO_3 + CO_2$ (1)

After calcination, the materials were homogenised using a Mortar Grinder (Pulverisette 2, Fritsch) for 60 seconds to eliminate any potential agglomeration. $Ca(OH)_2$ and Li_2CO_3 were also used as CO_2 sorption promoters (10 mol%) by addition after the calcination step.

107 The resulting sorbents were grinded and characterized by different techniques, including 108 powder X-ray diffraction (XRD), thermo-gravimetric analysis (TGA) and Fourier 109 Transformed Infrared (FTIR). A Bruker Nonius X8-Apex2 CCD diffractometer equipped 110 with an Oxford Cryosystems Cryostream, typically operating at 100 K was used for the XRD 111 analysis. A PerkinElmer Frontier infrared spectrometer was used to gain additional 112 information on the crystal structure of the synthetised sorbents before and after the CO_2 113 absorption experiments.

The CO₂ capture capacity of the resultant samples was measured by using a TGA (TA Q500). 114 About 15 mg of sample was loaded in the TGA pan for each experiment. Prior to CO₂ 115 sorption testing, the samples were dried in N₂ flow (95 mL/min) for 1 hr at the same 116 temperature used for the CO₂ sorption (600,700 °C). The heating rate used to rise the 117 temperature to 600 or 700°C (before holding it for 1hr) was 25 °C/min. Then, the CO₂ 118 absorption tests were performed by flowing 100% CO₂ gas at the desired temperature. The 119 120 weight increase due to CO₂ sorption (mmol CO₂/g sorbent) was measured as a function of time at a constant temperature (600, 700 °C) and constant concentration of CO₂ or flue gas 121 (95 mL/min) at atmospheric pressure. CO₂ uptake was calculated based on dry sorbents 122 weight (after drying step). Triplicate measurements were conducted to estimate the % error in 123 124 the experimental work. The measurements error (calculated as the standard deviation of triplicates) resulted lower than 5%. The CO₂ desorption step was carried out at the same 125 sorption temperature by switching the furnace atmosphere from CO₂ to N₂ for 1hr. Finally, 126 ten (five for the sorbent with 10% Li₂CO₃ additive) regeneration cycles were used to 127 ascertain the stability for selected sorbents. 128

129

130 **Results & Discussion**

131 Sorbents characterisation

132 The CO_2 capture reaction in presence of potassium silicate is described by the equation:

- 133 $K_2SiO_3 + CO_2 \rightarrow K_2CO_3 + SiO_2$
- 134

(2)

135 The reaction thermodynamics (Table 1) indicate that the capture reaction is overall 136 spontaneous ($\Delta G < 0$), and at temperatures < 100°C ΔS and ΔH are negative, indicating 137 exothermic reaction. However, similarly to other alkali silicates, this experimental work 138 clearly indicates that CO₂ capture is favoured at high temperature (Sanna et al., 2014).

139

| 140 | Table 1. | Thermodynamics of | of CO ₂ capture | e by K-FA iı | n gas phase a | t different temperatures |
|-----|----------|-------------------|----------------------------|--------------|---------------|--------------------------|
|-----|----------|-------------------|----------------------------|--------------|---------------|--------------------------|

| т, °С | deltaH, kJ | deltaS, J/°K | deltaG, kJ | К | Log(K) |
|-------|------------|--------------|------------|---------|--------|
| 0 | -1409.8 | -165.2 | -1364.7 | 1.E+261 | 261.0 |
| 100 | -1397.1 | -125.7 | -1350.2 | 1.E+189 | 189.0 |
| 200 | -1382.4 | -90.9 | -1339.4 | 8.E+147 | 147.9 |
| 300 | -1365.9 | -59.4 | -1331.9 | 2.E+121 | 121.4 |
| 400 | -1347.6 | -30.0 | -1327.4 | 1.E+103 | 103.0 |
| 500 | -1327.3 | -1.7 | -1326.0 | 4.E+89 | 89.6 |
| 600 | -1307.9 | 21.9 | -1327.0 | 2.E+79 | 79.4 |
| 700 | -1289.2 | 42.2 | -1330.2 | 3.E+71 | 71.4 |
| 800 | -1269.8 | 61.1 | -1335.4 | 1.E+65 | 65.0 |
| 900 | -1220.1 | 104.5 | -1342.6 | 6.E+59 | 59.8 |
| 1000 | -1198.0 | 122.5 | -1353.9 | 4.E+55 | 55.6 |

¹⁴¹ 142

The major chemical bonds present in the synthetised alkali metals silicate were identified by 143 FTIR spectra, as shown in Fig. 1. The characteristic absorption band from deformation of 144 M⁺OH become Si-O-M⁺ are shown at 880 and 994 cm⁻¹ (Hindryawati et al, 2014), while the 145 predominant absorbance peak at 1360 cm⁻¹ is due to siloxane bonds (Si-O-Si) (Thuadaij et al, 146 2008). The peak at 1427 cm⁻¹ can be attributed to CO_3^{2-} anion originating from the potassium 147 carbonate. As reported in previous work, the broadband between 2800 and 3500 cm⁻¹ is 148 attributed to silanol OH groups and adsorbed water (Kalapathy et al, 2000). The FTIR spectra 149 show formation of silicate structure. 150

Figure 2 shows the XRD patterns of the synthetised K-FA. It was found that the X-ray 151 diffraction patterns of the two materials contain identical peaks but with different intensities. 152 The major peaks were identified and compared to potassium metasilicate (K₂SiO₃), potassium 153 aluminium silicate $(K_{1,25}Al_{1,25}Si_{0,75}O_4)$ and potassium carbonate (K_2CO_3) using XRD 154 database and previous publication (N. Hindryawati et al, 2014). It was found that K-FA 1:1 155 156 structure presented a less defined crystal structure (see theta region between 15 and 25) and with the increase of K_2CO_3 used in the K_2CO_3 :FA mixture (2:1), the materials structure 157 become more crystalline. The intense diffraction sharp peaks from 12 to 43 are associated to 158 potassium metasilicate and potassium aluminium metasilicate, while the sharp diffraction 159 160 peaks at 12 and 26 are solely attributed to K₂SiO₃. As can be seen in Figure 1, K-FA 2:1

161 presents more intense peaks (1456 cm⁻¹) related to K_2CO_3 . This indicates a not complete 162 reaction during synthesis, probably due to poor contacts between K_2CO_3 and SiO₂ from FA.



167 Figure 2. XRD patterns of K-FA 1:1 and K-FA 2:1. Mineral phases identified: 1 Potassium aluminium silicate;

168 2 Potassium carbonate; 3 Potassium metasiilicate.

171 CO₂ capture studies

CO₂ capture experiments were carried out using the synthetized K-sorbents shown in Table 2. 172 The K-FA sorbents capacity was compared to those of Li-based and Li/Na-based fly ash 173 174 sorbents previously developed by our group (Olivares-Marin et al, 2010; Sanna et al, 2015). The sorption experiment at 500°C was used to compare the K-FA sorbent prepared here to 175 the Li-FA sorbent, while the sorption at 600°C was used to evaluate the differences with the 176 Li-Na-FA sorbents. Table 2 shows that the K-Fa 1:1 CO₂ sorption capacity was three times 177 higher than that of Li-FA under the same conditions, while K-FA 2:1 possesses a CO₂ 178 sorption capacity double than that of Li-FA. The K-FA materials had however a lower 179 180 capacity compared to the Li-Na-FA sorbent at both 500 and 600°C.

Table 2 also shows a clear difference in the CO_2 sorption capacity when different K_2CO_3 :FA molar ratio were used. The K-FA sorbent shows better performance at molar ratios 1:1 (0.32, 0.36 and 1.45 mmol CO_2 / g sorbent at 500, 600 and 700°C, respectively), while a K-FA ratio of 2:1 was able to chemisorb only 0.23, 0.36, 0.82 mmol CO_2 / g sorbent at the same temperatures. This can be related to the different distribution of mineral phases formed during the sorbent synthesis (see Figure 2), and a large presence of unreacted K_2CO_3 is present in the K-FA 2:1 (see Figure 1).

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Table 2. Summary of the CO₂ capture experiments in presence of 100%CO₂. 1:1 represents the K₂CO₃:SiO₂

190 (FA) molar ratio.

| | | | CO ₂ sorptio | n capacity |
|---|-----------|------------------------------------|----------------------------------|------------|
| | | | mmol CO ₂ / g sorbent | |
| | | Additive | | |
| Sample | Temp., °C | (mol%) | 1 h | 2h |
| Olivares-Marín et al., 2010 | | | | |
| $FA-Li_4SiO_4$ (1 SiO_2 : 2 Li_2CO_3) | 500 | | < 0.11 | na |
| FA-Li ₄ SiO ₄ (1 : 2) | 600 | 10% K ₂ CO ₃ | 0.59 | na |
| FA-Li ₄ SiO ₄ (1 : 2) | 600 | 20% K ₂ CO ₃ | 1.27 | na |
| FA-Li ₄ SiO ₄ (1 : 2) | 650 | 10% K ₂ CO ₃ | 0.48 | na |
| Sanna et al., 2015 | | | | |
| Li/Na-FA (0.5:0.5:1) | 500 | | 0.48 | |
| Li/Na-FA (1:1:1) | 600 | | 0.86 | 1 |
| This work | | | | |
| K-FA 1:1 | 500 | | 0.32 | 0.34 |
| K-FA 1:1 | 600 | | 0.36 | 0.39 |
| K-FA 2:1 | 500 | | 0.23 | 0.27 |
| K-FA 2:1 | 600 | | 0.36 | 0.45 |
| K-FA 2:1 | 700 | | 0.82 | 0.86 |
| K-FA 1:1 | 700 | | 1.45 | 1.50 |

192 Effect of temperature and CO₂ promoters

- Figures 3 and 4 show the effect of temperature on the CO₂ capacity. The initial weight 193 decrease was obtained in presence of N₂ flow, and represents desorption of absorbed water. 194 The CO₂ sorption experiments indicate that K-FA 1:1 and 2:1 increase their CO₂ sorption 195 capacity with temperature, with maximum sorption taking place at 700°C. This trend did not 196 differ from that previously found in presence of Li- and Li-Na- silicate sorbents (Sanna et al, 197 2015). As shown in Table 2, the K-Fa 2:1 CO₂ uptake (0.82 mmol CO₂/g sorbent after 1 hour 198 at 700°C) is about half compared to the CO₂ absorbed by using K-FA 1:1 at the same 199 200 temperature (1.45 mmol CO₂/g sorbent). Not only the CO₂ sorption capacity was higher at 700°C, but also the sorption rate was faster as indicated by the slope of the sorption profiles 201 in Figures 3 and 4. However, full desorption was not achieved under the studied conditions, 202 as CO₂ desorption did not occur with both K-FA 1:1 and 2:1 and even in presence of K-FA 203 doped with Ca(OH)₂. Also, sorption took longer to achieve maximum capacity compared to 204 using Li-FA sorbents under the same conditions, which results too long for industrial 205 applications (Sanna et al, 2015; Olivares-Marín et al, 2010). 206
- Since the pure K-FA silicates failed to desorb the CO_2 , a series of experiments were carried out to establish if the desorption could be improved by the presence of additives. Since previous works indicated that Li⁺ provides the highest CO₂ capture capacity among all the univalent cations, if provided in the same amount in the ceramic (Walton et al, 2006; Ridha et al, 2009), was selected Li₂CO₃ as sorbent promoter and Na₂CO₃ and Ca(OH)₂ for comparison.
- Figure 3 compares the CO₂ working capacity of K-FA 2:1 with and without Ca(OH)₂ at 213 700°C. The pure K-FA 2:1 sample captured 0.82 mmol CO₂/g. The CO₂ captured increased 214 to 1.04 mmol/g in presence of 10% Ca(OH)₂. Despite this, the Ca-additive was not able to 215 promote CO₂ desorption, indicating the formation of a products layer (carbonates and 216 oxides), which prevents CO₂ desorption. The addition of 10% Li₂CO₃ to K-FA 2:1, further 217 increased the CO₂ sorption to 2.25 mmol CO₂/g as can be seen in Figure 6-a. Therefore, the 218 219 presence of Li₂CO₃ promoted the CO₂ uptake and also the CO₂ sorption rate, as can be seen comparing the slopes in the sorption profiles in Figures 3 and 6-a. 220
- 221 Different additives were tested in the weight ratio of 10% using the K-FA 1:1 sorbent at

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water in N_2 required longer time, so that the real CO_2 working capacity is larger than shown

700°C, as presented in Figure 5. Figure 5-c indicates that the initial desorption of absorbed

in Figure 6-b. Both Na₂CO₃ and Li₂CO₃ were successful in desorbing all the CO₂ previously

absorbed. The CO₂ loading capacity was also enhanced to 2.39 mmol CO₂/g sorbent form 225 1.45 mmol/g in presence of Li₂CO₃. Moreover, fast sorption and desorption in presence of 226 10% Li₂CO₃ were achieved, requiring only minutes to fully uptake the CO₂ and about 20 227 minutes to desorbing it. The synergic effect of adding 10% Li₂CO₃ can be explained with full 228 229 activation of the Li-K eutectic melt at around 700°C, which increased CO₂ capacity and sorption rate (Sanna et al, 2015; Flores-Martínez and Pfeiffer, 2015). The CO₂ sorption 230 capacities in presence of K-FA 1:1 and 2:1 were comparable to those shown by Li-FA 231 sorbents (0.48 mmol CO₂/g sorbent) in presence of 10% K₂CO₃ additive at 650°C (Olivares-232 Marín et al, 2010). In the same study, 40% K₂CO₃ was required (at 600°C) to absorb 2.43 233 234 mmol CO₂/g sorbent.

Based on our observations and previous works (Ortiz-Landeros et al, 2012; Sanna et al, 2015; Flores-Martínez and Pfeiffer, 2015), the different CO₂ sorption capacities and rates observed at 500, 600 and 700°C can be explained by surface chemisorption limited by the formation of external shell containing an KAlO₂ secondary phase and Li₂CO₃ product at temperatures up to 500°C. Between 500 and 600°C, partial molten generated by eutectic phases starts to promote diffusion processes and then, at about 700°C, the CO₂ chemisorption is kinetically controlled by diffusion processes (Flores-Martínez and Pfeiffer, 2015; Sanna et al, 2015).

Moreover, it has been indicated that the binary compound $K_2O \cdot Al_2O_3$ (KAlO₂) shows that at 242 ambient pressure, KAlO₂ consists of tetragonal crystals between 1350 and 600°C (γ-KAlO₂) 243 and orthorhombic crystals below 600°C (β -KAlO₂), with the structural re-arrangement 244 comparable with that of NaAlO₂ and LiAlO₂ (de Kroon et al, 2001). Even if isolated silicates 245 present melting points well above 700°C, compounds with the composition Na_{1.55}K_{0.45}Si₂O₅ 246 247 have eutectic decomposition temperature at 696°C (Yazhenskikh et al, 2008). The Li cations can decrease the melting point of the sorbent enhancing the diffusion of the species on the 248 which results in enhanced CO_2 sorption capacity and CO_2 249 sorbent surface, sorption/desorption rates. Moreover, the formed eutectic phases in presence of Li₂CO₃ clearly 250 increase the carbonate decomposition processes by enhancing the diffusion of cations and 251 CO₂, enabling the recyclability of the material, as shown in Figure 5 (Flores-Martínez and 252 253 Pfeiffer, 2015).

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259 Figure 3. TGA of (a) K-FA 2:1 700°C+10% Ca(OH)₂; (b) K-FA 2:1700°C; (c) K-FA 2:1 600°C; (d) K-FA 2:1
260 500°C.



Figure 4. TGA of (a) K-FA 1:1 700°C; (b) K-FA 1:1 600°C; (c) K-FA 1:1 500°C.



264

265 Figure 5. TGA of (a) K-FA 1:1; (b) K-FA 1:1+10% Na₂CO₃; (c) K-FA 1:1+10% Li₂CO₃.

267 Sorbents regeneration

Figure 6 shows 10 and 6 CO₂ sorption/desorption cycles for the Na-FA sorbent with 268 K₂CO₂:SiO₂ molar rations of 2:1 and 1:1 in presence of 10% Li₂CO₃. Both K-FA 2:1 and 1:1 269 were able to maintain 100% CO₂ sorption capacity at 700°C, respectively after 10 and 6 270 cycles. K-FA 1:1 showed a slightly increase of the CO₂ sorption cycle after cycle, which 271 could be ascribed to the not completed initial desorption under N2 atmosphere before the 272 cyclic CO₂ sorption/desorption. K-FA presented superior cyclic stability compared to other 273 274 high temperature absorbents, such as hydrotalcites, where the capacity of the latter decrease after a number of cycles (Maroño et al, 2014). 275

K-Fa 1:1 with 10% Li₂CO₃ after 6 sorption/desorption cycles was analyzed by XRD to 276 establish any structural change in relation to the increased CO₂ uptake capacity shown after 6 277 cycles. The XRD patterns of the starting K-FA 1:1 and the K-FA after 6 cycles are shown in 278 Figure 7, which indicates that, the sorption/desorption cycles affects the mineral composition 279 280 of these materials. In fact, there was an increase of sodium aluminium silicate phase and a decrease of the abundance of metasilicate phases. This behavior was already observed in our 281 previous work using Li-Na-FA silicates (Sanna et al, 2015). Both TGA and XRD data 282 suggest that the crystals structure modification improves the CO₂ sorption, which may be 283

related to enhanced K species diffusion due phase change (from solid to liquid-like) of
eutectic Na-meta-silicates phases in presence of impurities such as K, Ca and Mg oxides
(from fly ash) as indicated in previous work (Sanna et al, 2015; Flores-Martínez and Pfeiffer,
2015).











Figure 7. XRD pattern of raw K-FA and K-FA after 6 absorption/desorption cycles. 1 Potassium aluminium
 silicate (K_{1.25}Al_{1.25}Si_{0.75}O₄); 2 Potassium carbonate; 3 Potassium metasiilicate.

295 Presence of diluted CO₂ and potential scale up

The use of solid sorbents for CO_2 capture implies that the selected materials present rapid CO₂ sorption/desorption rates and high CO₂ sorption capacity. This is required due to the diluted CO₂ in flue gas, which results in large volumes to be treated.

299 The effectiveness of the K-FA sorbents in capturing CO₂ from a diluted stream was investigated. Figures 8 and 9 show the variation of the CO₂ working capacity for the K-FA 300 2:1 and K-FA 1:1 in presence or absence of 10% Li₂CO₃. The CO₂ uptake with diluted CO₂ 301 (14%) were lower than those obtained in presence of pure CO₂ at 500°C, similar at 600°C and 302 somehow higher at 700°C (1.08 mmol CO₂ / g sorbent vs 0.82 for K-FA 2:1 and 1.50 vs 1.45 303 for K-FA 1:1) indicating that the eutectic melt favours CO₂ diffusion in/out from the sorbent 304 surface. However, the slopes of the sorption and desorption profiles suggest lower reaction 305 306 rates compared with those in presence of pure CO₂ stream and also, desorption was not completed in reasonable times. Diluted CO₂ resulted in a lower CO₂ sorption when the K-FA 307 1:1 was tested in presence of 10% Li₂CO₃ (1.84 mmol CO₂ / g sorbent vs 2.39), but reactions 308

- 309 occurred faster and desorption was completed, compared to the tests carried out in absence of310 additive.
- 311 A 500 MW coal-fired power plant with 90% CO₂ capture (~419,000 kg CO₂/h) was
- 312 considered to evaluate the potential industrial scale application of the developed sorbent.



314 Figure 8. K-FA 2:1 CO₂ sorption at a) 500°C, b) 600°C and c) 700°C in presence of 14%CO₂.



- **Figure 9.** CO₂ sorption of K-FA 1:1. a) 100%CO₂ at 500°C; b) 100%CO₂ at 600°C; c) 100%CO₂ at 700°C; d)
- 317 14% CO₂ at 700°C in presence of 10% Li₂CO₃; e) 100% CO₂ at 700°C in presence of 10% Li₂CO₃.
- 318
- The capture plant was sized considering fix bed units able to load a total of 28000 kg of sorbent, as in previous work, to evaluate the feasibility of the developed sorbents (Quinn et al, 2012).
- 322 K-FA 1:1 10% Li₂CO₃ was selected for the calculations due to its high total CO₂ capacity of
- 323 2.4 mmol CO_2/g sorbent after 1 hour at 700°C, as can be seen in Table 3. Also, a flue gas 324 containing 14%CO₂ (balanced with N₂) was considered.
- Four hypothetical cases were considered with adsorption taking place in 60, 30, 15 and 5 325 minutes to show how uptake time affects the hypothetical plant footprint. The number of 326 327 reactors required to capture 90% of the CO₂ emitted in a 500 MW coal power plant is 328 summarised in Table 3. Since the calculated number of vessels based on 60 minutes were 329 very large (3.9 mmol CO₂/g sorbent for a total of 140 vessels), they were not reported in the table. However, taking advantage of the rapid sorption and desorption rates of the K-FA 330 331 sorbent in presence of the Li₂CO₃ sorption promoter (10 wt%), the process became feasible at 5 minutes sorption time in presence of pure CO_2 stream, with only 12 vessels in parallel 332 required for the CO₂ uptake, which would result in reduced capital costs. As expected, the 333 CO₂ sorption capacity in presence of diluted CO₂ stream (14% CO₂ balanced with N₂) is 334 lower than in presence of pure CO_2 and this affects the foot-print of the hypothetical CO_2 335 capture plant, where 20 reactors would be required. 336
- 337
- **Table 3.** CO₂ sorption capacity at different times and number of reactors required to capture 90% of CO₂ from a
- 500 MW coal-fired power plant with 90% CO₂ capture (~ 419,000 kg CO₂/h). na: non available.

| | CO ₂ sorption, mmol CO ₂ /g sorbent | | | | |
|--------------------------------|---|----------------|------------------------------|---------------|--|
| Time, min | 5 | 10 | 15 | 30 | |
| K-FA 1:1 | 2.38 | 2.4 | 2.4 | 2.4 | |
| K-FA 2:1 | 1.44 | 1.58 | 1.67 | 1.78 | |
| | Reactors required for CO ₂ sorption | | | | |
| | Reactors | required | for CO ₂ so | orption | |
| K-FA 1:1 - 100%CO ₂ | Reactors 12 | required 24 | for CO ₂ so 35 | orption 71 | |

³⁴⁰ 341

Amount of sorbents can be lowered by employing rapid thermal swing chemisorption (RTSC) process that utilizes shell and tube type vessels to minimize heating and cooling time (Lee and Sircar, 2008). Calculations indicated a 24% reduction in capital cost and a 78% reduction in operating cost relative to conventional MEA technology may be achievable (Quinn et al, 2012). For example, the K-FA sorbent able to load 2.38 mmol CO_2/g sorbent would result in 0.37 t to be loaded in the RTSC. Further works are required to evaluate the potential use of these sorbents with rapid sorption/desorption processes in presence of moisture.

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351 Conclusions

Potassium-fly ash (K-FA) CO₂ chemisorption at high temperature was evaluated using 352 353 isothermal and cyclic thermo-gravimetric analyses in presence of 100% and 14% CO₂. Lithium carbonate and calcium hydroxide were also evaluated as CO₂ sorption promoters. 354 Temperature strongly affected the performance of the K-FA sorbents, resulting in a CO₂ 355 uptake of 1.45 mmol CO₂/g sorbent for K-FA 1:1 at 700°C. The CO₂ sorption was enhanced 356 by the presence of lithium carbonate (10 wt%), with the K-FA 1:1 capturing 2.38 mmol 357 358 CO₂/g sorbent at 700°C in 5 minutes. The presence of 10% Li₂CO₃ also accelerated sorption and desorption. The results suggest that the increased uptake of CO₂ and faster reaction rates 359 in presence of K-FA can be ascribed to the formation of K-Li eutectic phase, which favours 360 361 the diffusion of potassium and CO₂ in the material matrix. The cyclic experiments showed that the K-FA materials maintained stable CO₂ uptake and reaction rates over 10 cycles. 362 363 Thanks to the rapid CO_2 sorption/desorption of the K-FA sorbent with Li_2CO_3 (10 wt%), the process would require only 12 fix bed units for the sorption of 90% of the CO₂ emitted in a 364 365 500 MW coal power plant. As expected, the CO₂ sorption capacity in presence of 14% CO₂ affects the foot-print of the hypothetical CO₂ capture plant, where 20 reactors would be 366 367 required.

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