

1 **Phytoremediation: Green technology for the removal of mixed**
2 **contaminants of a water supply reservoir**

3
4 **Sabrina Loise de Moraes Calado¹, Maranda Esterhuizen-Londt², Helena Cristina Silva**
5 **de Assis¹, and Stephan Pflugmacher^{2,3*}**

6 ¹Department of Pharmacology, Federal University of Paraná - Avenue Coronel Francisco Heráclito dos Santos,
7 210, Jardim das Américas, Curitiba, Paraná, Brazil - CEP: 81531-990; ²Ecotoxicology in an Urban
8 Environment, Ecosystems and Environmental Research Programme, Faculty of Biological and Environmental
9 Sciences, University of Helsinki, Niemenkatu 73, 15140 Lahti, Finland; ³Korea Institute of Science and
10 Technology Europe (KIST), Joint Laboratory of Applied Ecotoxicology, Campus 7.1, Saarbrücken, Germany

11
12 Sabrina Calado: sahbio@hotmail.com

13 Maranda Esterhuizen-Londt: maranda.esterhuizen-londt@helsinki.fi (ORCID: [https://orcid.org/0000-](https://orcid.org/0000-0002-2342-39419)
14 [0002-2342-39419](https://orcid.org/0002-2342-39419))

15 Helena Cristina Silva de Assis: helassis@ufpr.br

16 *Corresponding author: Stephan Pflugmacher: stephan.pflugmacher@helsinki.fi (ORCID:
17 <https://orcid.org/0000-0003-1052-2905>)

18
19
20 *Contributions:* SC: Experiment execution, data analysis, manuscript preparation; MEL: LC-MSMS method
21 development and analysis, experimental planning and supervision, manuscript preparation; HCSA:
22 manuscript preparation and research supervision; SP: Research concept development, research supervision.

23 **Abstract**

24 The Iraí Reservoir, a water supply in Brazil, is constantly impacted by
25 anthropogenic activities such as waste inputs from agriculture, hospitals, and
26 urbanization, resulting toxic cyanobacterial blooms causing economic, social, and
27 environmental problems. The present study assessed the concentration of some
28 common contaminants of the Iraí Reservoir, namely paracetamol, diclofenac, and
29 microcystin-LR and tested whether a laboratory scale Green Liver System® would
30 serve as a suitable technology to remove these contaminants. Further, the study
31 investigated whether the pollutants caused adverse effects to the macrophytes using
32 catalase as a biomarker for oxidative stress and investigated whether
33 biotransformation (glutathione S-transferase) was a main route for detoxification.
34 *Egeria densa*, *Ceratophyllum demersum*, and *Myriophyllum aquaticum* were
35 exposed to a mixture of the three contaminants for 14 days in a concentration range
36 similar to those detected in the reservoir. The plants removed 93 % of diclofenac
37 and 100 % of MC-LR after 14 days. Paracetamol could not be detected. Catalase
38 and glutathione S-transferase enzyme activities remained unaltered after the 14-day
39 exposure, indicating that the mixture did not cause oxidative stress. The study
40 showed that the aquatic macrophytes used are suitable tools to apply in a Green
41 Liver System® for the remediation of mixed pollutants.

42

43 **Keywords:** Microcystin-LR, diclofenac, paracetamol, Green Liver System®,
44 phytoremediation, aquatic macrophytes.

45

46 **1. Introduction**

47 Aquatic ecosystems are continuously affected by anthropogenic activities such as
48 nitrogen and phosphorous inputs, mainly from agriculture, as well as other toxic compounds.
49 These compounds can change the ecosystem dynamics, cause eutrophication, possibly resulting
50 in cyanobacterial blooms which release toxins, all affecting the aquatic biota (Schulz et al. 2015;
51 Scholz et al. 2017). Moreover, water contamination can result in human exposure via food and
52 drink, thereby posing a threat to human health (Gibble et al. 2016). Water supply reservoirs are
53 also affected by contamination resulting in high costs to the water treatment facilities to ensure
54 safe drinking water (Calado et al. 2017).

55 The most commonly occurring cyanobacteria *Microcystis aeruginosa* is known to
56 produce hepatotoxic microcystins (MCs), especially microcystin-LR (MC-LR) (Gupta et al.
57 2003; Omidia et al. 2018). Several studies have reported on the toxic effects of the MC-LR
58 including liver failure in humans (Yuan et al. 2006), oxidative stress in aquatic organisms
59 (Amado and Monserrat 2010), and molecular damage in mammals (Zegura et al. 2011). In
60 1998, the World Health Organization established the limit for MC-LR (1000 ng/L) in drinking
61 water (WHO 1998) and in Brazil the monitoring of cyanobacteria and cyanotoxins for water
62 control was incorporated in 2000 (Brazil 2011). MCs are common in Brazilian reservoirs and
63 studies have reported concentrations ranging from 0.5 to 4.5 µg/L (Fernandes et al. 2005;
64 Ferrão-Filho et al. 2014; Hauser-Davis et al. 2015).

65 Due to wastage and only partial adsorption and metabolism of pharmaceuticals in
66 humans, high concentrations of many pharmaceuticals have been detected in aquatic
67 environments. Some pharmaceuticals cannot be completely removed by the current
68 conventional water treatment processes and the population ingests these compounds via
69 drinking water on a daily basis (Lonappan et al. 2016). Pharmaceuticals such as diclofenac,
70 paracetamol, ibuprofen, and penicillin have been found in aquatic environments such as ground,
71 and drinking water (Ebele et al. 2017; Yang et al. 2017).

72 Studies have reported that paracetamol can cause toxic effects in low concentrations
73 (Nunes et al. 2014) and the effects in aquatic organisms have been reported (Guiloski et al.
74 (2017b). Similarly, diclofenac is commonly found in aquatic ecosystems with numerous authors
75 reporting accumulation in aquatic organisms causing damage (Cunha et al. 2017; Näslund et al.
76 2017, Liu et al. 2017; Gröner et al. 2017; Guiloski et al. 2017a). For this reason, the European
77 Commission established the maximum allowed limit of 100 ng/L in drinking water and has
78 declared diclofenac as a hazardous substance (European Commission 2012).

79 The Iraí Reservoir is located in the South of Brazil and is used as a potable water supply.
80 There are many anthropogenic activities occurring around the reservoir such as agriculture,
81 industries, hospitals, and settlements causing contamination and eutrophication leading to
82 frequent cyanobacterial blooms. For this reason, several pharmaceutical, agricultural, and other
83 chemical contaminants, including cyanobacterial toxins have been found in this water body
84 (Bittencourt-Oliveira 2003; Kramer et al. 2015) like many others in the region. Due to the
85 occurrence of these compounds in many reservoirs in developing countries, there a is a need
86 for a low cost, sustainable, easy to manage, eco-friendly remediation technique ensuring safe
87 drinking water.

88 Phytoremediation is a green technology used as a tool to improve and complement the
89 water treatment processes. The Green Liver System® was recently reported as a suitable,
90 sustainable, and environmentally friendly approach to remediate contaminated water bodies
91 (Pflugmacher et al. 2015) showing success in the remediation of pharmaceuticals (Vilvert et al.
92 2017) and cyanobacterial toxins (Pflugmacher et al. 2016). It is a methodology that purifies
93 water in a short time frame, using the uptake and biotransformation capacity of aquatic
94 macrophytes. “Green Liver” in the name refers to the fact that the plants work as an animal liver
95 in the biotransformation and detoxification of compounds (as detailed in Pflugmacher et al.
96 2015). However, to achieve efficient water purification it is necessary to replace the plants to
97 avoid release of the metabolites (Pflugmacher et al. 2015).

98 In addition to evaluating the uptake of contaminants by plants, the effects on the plants
99 also need to be assessed as mortality of the plants could lead to release of the contaminants.
100 Several studies have investigated the activities of antioxidative enzymes, e.g. catalase (CAT),
101 superoxide dismutase (SOD), glutathione reductase (GR) and glutathione peroxidase (GPx),
102 and biotransformation, as the activity of glutathione *S*-transferase (GST), as physiological
103 biomarkers for adverse effects (Pflugmacher et al. 2007; Flores-Rojas et al. 2015; Spengler et
104 al. 2017). These enzymes prevent cellular damage by degrading reactive oxygen species (ROS)
105 such as O₂•, H₂O₂ and OH•, which are localized in chloroplast, mitochondria, and peroxisomes
106 (Gill and Tuteja 2010). ROS can be produced under normal cellular metabolism or produced
107 from xenobiotic exposure (Fernández-Fuego et al. 2017). GST, a biotransformation enzyme,
108 conjugates electrophilic compound with glutathione (GSH), playing a role in the defense against
109 oxidative damage (Van der Oost et al. 2003). Several studies have used this biomarker to
110 evaluate the environmental stress in invertebrates, vertebrates, and plants (Pradhan et al. 2016;
111 Lajayer et al. 2017). The use of these biomarkers is an advantageous tool in ecotoxicological
112 studies and phytoremediation programs.

113 Based on the need for an eco-friendly, cost efficient remediation technology for
114 developing countries i.e. to address the water quality issues of Brazil and based on the previous
115 success reported with the Green Liver System® (Pflugmacher et al. 2015), the aims of the
116 present study were therefore to (1) assess the concentration of the three drinking water
117 contaminants of emerging concern in the Iraí Reservoir, Brazil, i.e. paracetamol (690 ng/L),
118 diclofenac (12500 ng/L), and MC-LR (2030 ng/L), and to (2) evaluate the efficiency of the
119 Green Liver System® to remove these three contaminants in the concentrations that were found
120 in the reservoir, using *Egeria densa*, *Ceratophyllum demersum* and *Myriophyllum aquaticum*
121 and to (3) assess the macrophytes' physiological responses to the exposure by monitoring the
122 enzyme activities of CAT and GST.

123 **2. Material and methods**

124 ***2.1 Sampling from the Iraí Reservoir***

125 Water samples were collected from the Iraí Reservoir on March 2017 using dark bottles
126 with a total volume of 1 L according to Pierre Gy's theory of sampling principles (Pitard 1993).
127 The water samples were frozen, concentrated by lyophilization (-48.3 °C, 0.1163 mbar), and
128 resuspended in 70 % methanol before quantification via liquid chromatography tandem mass
129 spectroscopy (LC-MS/MS). The yield of the compounds after lyophilization was evaluated
130 before sample treatment. Three emerging contaminants, namely paracetamol, diclofenac, and
131 MC-LR, were analyzed.

132 In short, acetaminophen (paracetamol) was quantified according to Esterhuizen-Londt
133 et al. (2016), diclofenac was quantified according to Esterhuizen-Londt et al. (2017), and MC-
134 LR was quantified according to Balsano et al. (2015).

135 ***2.2 Recovery after lyophilization procedure***

136 Due to the low concentrations of the contaminants, an experiment to evaluate the
137 compounds lost in the lyophilization procedure was carried out. It was tested using a control in
138 Provasoli medium (Nimptsch et al. 2008) and a known concentration of the compounds
139 (paracetamol: 690 ng/L, diclofenac; 12500 ng/L and MC-LR: 2030 ng/L). After total
140 homogenization, water samples were frozen in liquid nitrogen, lyophilized, and re-diluted
141 methanol (MS grade) followed by analyzed on LC-MS/MS. For the exposure experiments, it
142 was decided to use a concentration 10 times higher than that quantified in Iraí Reservoir because
143 of the percentage loss due to lyophilization and according to concentrations in Iraí Reservoir
144 described in other studies.

145 ***2.3 Plant Material and chemicals***

146 *E. densa*, *C. demersum*, and *M. aquaticum* were purchased from ExtraPlant (Extragroup
147 GmbH, Münster, Germany). Aquatic macrophytes were maintained in tanks (100 L) over 7 days

148 for acclimation under controlled conditions, i.e. in pH 8 Provasoli media (Nimptsch et al. 2008),
149 at 20 ± 1 °C, and a photoperiod of 14 h light/10 h dark.

150 MC-LR was purchased from Alexxis GmbH (Grünberg, Germany). All other chemicals
151 were purchased from Sigma-Aldrich Chemie GmbH (Munich, Germany) unless stated
152 otherwise.

153 **2.4 Exposure experiments**

154 Exposure experiments were carried out in a model Green Liver System®. The system
155 was built using glass with a total volume of 50 L. This system was divided into 6 compartments,
156 which are constructed to allow the continuous flow of water through the system via pumping
157 (as depicted in Nimptsch et al. 2008). *E. densa* was added to the first and second compartments
158 (n=4), *C. demersum* was added to the third and fourth compartments (n=4) and *M. aquaticum*
159 was added to the fifth and sixth compartments (n=4). Each plant species, which was fully
160 grown, had a mass of circa 150 g. Plants were exposed to 690 ng/L paracetamol, 12500 ng/L
161 diclofenac and 2030 ng/L MC-LR. The negative control consisted of the same conditions
162 without the addition of the compounds. The positive control consisted of running the system
163 with the contaminant mixture without plants. The experiment was performed during 14 days
164 and media samples were collected on day 0, 1, 3, 7 and 14. On day 14, all the plants samples
165 were collected. Five replicates were carried out for water and plant samples. During the
166 exposure, the Green Liver System® was kept under the same conditions as during
167 acclimatization.

168 The water samples collected were frozen in liquid nitrogen and lyophilized as stated
169 before. Afterwards, the samples were resuspended in 1 mL of methanol (MS-grade). The
170 samples were stored at -20 °C until LC-MS/MS analysis.

171 **2.5 Extraction procedure and quantitative analysis**

172 Plants samples were collected on the end of the experiment in order to analyze the uptake
173 of the plants. Samples were frozen in liquid nitrogen and ground to a fine powder. Samples (0.1

174 g) were then added to MS-H₂O for paracetamol and 70 % methanol for diclofenac and MC-LR
175 extraction and left shaking for 30 min before centrifugation at 3400 × g for 10 min. The
176 supernatant was collected and the pellet were washed in an equal volume of MS-H₂O, followed
177 by vortexing and centrifugation at 3400 × g for 10 min. The supernatants were pooled and
178 filtered using 0.45 µm syringe cellulose acetate filters. The samples were stored at -20 °C until
179 the LC-MS/MS analysis. Both the media and the extracted samples were analyzed on LC-
180 MS/MS as stated in section 2.1.

181 ***2.6 Enzyme activities***

182 The activities of CAT (EC 1.11.1.6) and GST (EC 2.5.1.18) were analyzed in order to
183 assess the oxidative stress status and biotransformation in the plant tissues after exposure.

184 The enzymes were extracted as detailed by Pflugmacher (2004). In short, the plant
185 samples were ground in liquid nitrogen and 1.5 g of the powder was suspended in 3 mL of 0.1
186 mol/L sodium phosphate (NAP) buffer (pH 6.5) containing 1 mmol/L EDTA, 20 % (v/v)
187 glycerol, and 1.4 mmol/L dithioerythriol. The samples were stirred for 20 min on ice before
188 centrifugation at 5400 × g for 10 min at 4 °C. After a second centrifugation step (86900 × g, 60
189 min), the microsomal pellet was resuspended in 0.5 mL of 20 mmol/L NAP buffer containing
190 20 % glycerol. The 35 to 80 % saturation fraction was collected by NH₄SO₄ precipitation,
191 stirring for 20 min, followed by centrifugation at 48900 × g for 30 min. The supernatant was
192 discarded and the pellet was dissolved in 1 mL of 20 mmol/L NAP (pH 7.0). The samples
193 desalted on NAP 10 column (GE Healthcare Life, Freiburg Germany). Protein determination
194 was spectrophotometrically performed according to Bradford (1976), using Bradford's reagent.

195 CAT activity was measured according to Baudhuin et al. (1964) measuring the
196 breakdown of H₂O₂ (200 mmol/L) as substrate to H₂O and O₂ at 240 nm. GST activity was
197 evaluated according to Habig et al. (1974), measuring the conjugation of CDNB (1-chloro-2,4-
198 dinitrobenzen) and GSH (60 mmol/L) at 340 nm.

199 ***2.7 Statistical analysis***

200 The statistical analysis was performed using the R software 3.2.2 in order to compare
201 the enzyme activities between the control and treatment plants. Levene's homogeneity test and
202 Shapiro-Wilk normality preceded the data analysis. T-test was used to analyze the statistical
203 differences between control and treatment plants. The significance level was $p < 0.05$.
204

205 **3. Results**

206 Paracetamol, diclofenac, and MC-LR were quantified in the reservoir water samples at
207 the concentrations of 69 ng/L, 1250 ng/L, and 203 ng/L respectively.

208 Lyophilization was used in the present study as a means to analyze the low
209 concentrations of the compounds present in Iraí Reservoir, which were below the lower limits
210 of detection of the LC-MS/MS methods. However, our results showed low recovery of the
211 compounds. The percentages of the recovery were: paracetamol <10 %, diclofenac 44.8 %, and
212 MC-LR 9.0 %. Therefore, the use of a 10-fold higher exposure concentration for the laboratory
213 experiments was selected.

214 Neither the water nor the plants of the negative control samples contained the tested
215 compounds. During the two-week exposure period, the aquatic macrophytes did not showed
216 any visible morphological alterations for neither the control nor the exposure sets.

217 **3.1 Paracetamol**

218 Due to the quantification limit of the method, it was not possible to quantify paracetamol
219 in water samples. However, paracetamol in *M. aquaticum* was well measurable after 14 days
220 (137.6 ± 5.1 ng/g FW). The total amount taken up by the plants was 41.0 % of the exposure
221 concentration of 690 ng/L.

222 **3.2 Diclofenac**

223 For the treatment samples, there was no significant decrease of the diclofenac
224 concentration in water samples after 7 days (Fig. 1). However, the diclofenac concentration in
225 the water samples decreased by 93.0 % after 14 days. When taking into account this result
226 compared to the control without plants, for which the degradation was 43.0 % in 14 days, this
227 is a significant reduction ($p < 0.05$). Diclofenac was measurable in *E. densa* (132.6 ± 30.1 ng/g
228 FW; 3.4 %) and *C. demersum* (160 ± 15.9 ng/g FW, 5.5 %). Of the total amount of diclofenac
229 (12500 ng/L), 8.9 % (1112.5 ng/L) was taken up by the macrophyte (Fig. 1).

230 **Figure 1 here.**

231 **3.3 Microcystin-LR**

232 The MC-LR concentrations in water samples from the treatment set decreased by 69 %
233 within 24 h and 100 % after 3 days (Fig. 2). In the parallel experiment without plants, the
234 degradation was 55 % after 7 days and 61 % after 14 days. The concentrations in plants were
235 below of the quantification limit (Fig. 2).

236 **Figure 2 here.**

237 **3.4 Enzyme activities**

238 CAT and GST activities were not significantly different between control and treatments
239 in *E. densa* ($t=0.2135$, $p=0.8414$; $M=4$, $p=1$), *C. demersum* ($t=-18883$, $p=0.1321$; $t=0.5125$,
240 $p=0.6363$), or *M. aquaticum* ($t=1.5797$, $p=0.1888$; $t=-0.2736$, $p=0.7979$) (Fig. 3).

241 **Figure 3 here.**

242 **4. Discussion**

243 The results of the chemical analysis showed that Iraí Reservoir is contaminated with a
244 mixture of compounds including diclofenac, paracetamol, and MC-LR. The concentrations of
245 these compounds can be associated with the anthropogenic activities around this water body.
246 Although this reservoir is used as a water supply, there are several anthropogenic activities that
247 contribute to input of pharmaceuticals and cyanotoxins in the water. The hospital and
248 settlements contribute to inputs of the pharmaceuticals in water (Santos et al. 2010), and the
249 agriculture activities contribute to the increase of nutrients and organic matter result in
250 cyanobacterial blooms (Scholz et al. 2017).

251 The measured MC-LR concentration (203 ng/L) was below of the legislation limit for
252 drinking water (1000 ng/L) (Brazil 2011); however, the diclofenac concentration (1250 ng/L)
253 was ten times more than that allowed by the legislation for drinking water (100 ng/L) (European
254 Commission 2012). Paracetamol is a pharmaceutical that is not incorporated in Brazilian or
255 European legislation. Since pharmaceuticals are present in water bodies at high concentrations,
256 it is important to regulate these emerging contaminants. These contaminants may cause

257 problems to the aquatic organisms and human health, and are particularly worrisome since they
258 cannot be totally removed by conventional water treatment (Lonappan et al. 2016; Guiloski et
259 al. 2017b).

260 Although minute concentrations of these compounds can pose a risk to the environment,
261 these low amounts are difficult to quantify. For developing countries restricted by their financial
262 dispositions, it is not possible to use state of the art, highly efficient methods as they are often
263 very expensive to implement. The lyophilization, selected as it is an inexpensive method,
264 proved to be inefficient to concentrate environmental concentrations to those analyzable on LC-
265 MS/MS. Furthermore, it means that, the concentration found in Iraí Reservoir can be
266 underestimated. Fonte et al. (2006) assessed the lyophilization procedure and reported that the
267 processing conditions can result in freezing and desiccation stress causing damage and
268 instability to the substances. In addition, any change in the process can transform an efficient
269 into inefficient process. In future studies, other methodologies to improve the recovery to the
270 quantification of compounds should be tested.

271 In the Green Liver System® experiment, paracetamol could not be quantified in water
272 samples, however, 41 % of the total amount of paracetamol could be quantified intracellularly
273 in *M. aquaticum* suggesting uptake. The other 59 % could have been biotransformed by the
274 plants, and/or by natural/bacterial degradation, surface bound or have not been up taken or
275 degraded. Paracetamol transformation can form metabolites such as N-acetyl-benzoquinomine,
276 glucuronide, sulfate, and mercapturate; and the biotransformation by plants can be via
277 glucoronisation and generation of conjugates with glutathione (Huber et al. 2009). Another
278 study that tested plants to remove paracetamol, but using wetlands, showed that paracetamol
279 was removed, however, it was attributed more to the degradation associated to the biofilm in
280 roots (Ranieri et al. 2011).

281 Diclofenac was reduced by 93 % in water samples and only 43 % for natural and/or
282 bacterial degradation after 14 days, suggesting that the plants are taking up this compound. The

283 results obtained are comparable to those achieved by Matamoros et al. (2012) within the same
284 time frame using only *C. demersum*. However, diclofenac concentrations in *E. densa* and *C.*
285 *demersum* tissues were only 8.9 % of the total amount and this result can be possibly attributed
286 to the biotransformation of diclofenac intracellularly.

287 Diclofenac can be transformed into several products such as diclofenac-lactam, 4'-
288 hydroxy-diclofenac, 5'-hydroxy-diclofenac, and diclofenac-benzonic acid. Studies have
289 suggested that transformation and biotransformation can occur via monooxygenation,
290 oxidation, decarboxylation, conjugation, and hydroxylation (Jewell et al. 2016; Bouju et al.
291 2016) and that the process can occur very rapidly; for example, Huber et al. (2012) reported
292 biotransformation after 3 h. Diclofenac and its metabolites were quantified in the plant tissues;
293 and after 7 days 66 % of diclofenac concentrations decreased in the plants. They also suggest
294 the biotransformation of diclofenac is via hydroxylation to 4'-hydroxy-diclofenac and
295 conjugation to glucopyranoside.

296 For the treatment samples, after 3 days in the presence of the plants, no MC-LR could
297 be detected, and in the control experiment the concentration remained stable for the first 3 days.
298 This result suggested that the plants took up MC-LR and can be used as tools in the pretreatment
299 of water from reservoirs that are constantly contaminated with MCs. The MC-LR
300 concentrations in plants were below of the quantification limit. However, other studies that used
301 5 to 10 times higher concentrations showed that MC-LR was up taken by the aquatic plants and
302 *C. demersum* was a successful plant to remove this toxin (Pflugmacher et al. 2015; Contardo-
303 Jara et al. 2015). In a study by Romero-Oliva et al. (2015) it was shown that *E. densa* had a
304 higher MCs bioaccumulation capability compared by *C. demersum*. The low concentrations
305 used in the present study may have resulted in low measurement in the plant tissue. In addition,
306 studies described that when MC-LR enter the cell, it binds to GSH and phosphatase proteins
307 (Bittencourt-Oliveira et al. 2013; Liu and Sun 2015) and the method used only quantifies free

308 MC-LR. However, it should be considered that after 14 days the MC-LR could have been
309 biotransformed in the plants.

310 *E. densa*, *C. demersum*, and *M. aquaticum* took up the composts differently. Diclofenac
311 was measured in *E. densa* and *C. demersum*, yet paracetamol could be only measured
312 intracellularly in *M. aquaticum*. This result showed the importance of working with different
313 plant species in phytoremediation programs, mainly in aquatic environments that are
314 contaminated by a mixture of compounds. In this context, different plants species respond in
315 different ways to the contaminants and the choice of the species used in phytoremediation is a
316 very important step.

317 Studies that evaluate the stress in plants, report the sensitivity of these organisms to
318 contaminants and it can determine which species can be used as bioindicators or for
319 phytoremediation. GST and cytochrome P450 enzymes can participate in the biotransformation
320 of paracetamol, diclofenac, and MC-LR. However, in the present study, the GST activities of
321 the exposed plants were not elevated when compared to the control. It can be due to analysis
322 only being conducted after 14 days and this enzyme's activity could have returned to normal
323 levels. Pflugmacher (2004) showed that in *C. demersum*, GST levels started to reduce after 48
324 h of exposure to 0.5 ng/L Mc-LR. Nunes et al. (2017), in a study with clams, reported that the
325 GST enzyme was activated in the first 96 h of the paracetamol exposure. After 10 days the GST
326 activity was similar among control and treatment groups. In addition, other biotransformation
327 enzymes that were not checked in the present study can be metabolizing these compounds such
328 as P450 monooxygenases and glycosyltransferases (Huber et al. 2012). Another study that
329 evaluated biotransformation of endocrine disrupting chemicals in *C. demersum*, suggested high
330 efficiency in the peroxidases metabolism for detoxification when compared with the GST
331 metabolism (Reis et al. 2014).

332 In the present study, CAT activity also was not statistically different between control
333 and treatment plants. This result suggests that the CAT activity was not elevated at all or

334 returned to normal after 14 days in the plant species at the concentrations used in this study.
335 Studies using *Hydrilla verticillata* (Spengler et al. 2017) and *Fucus vesiculosus* (Pflugmacher
336 et al. 2007) have reported the increase of the CAT activity after exposure to contaminants. In
337 addition, Kummerová et al. (2016), evaluating paracetamol and diclofenac exposure in *Lemna*
338 *minor* showed induction of stress oxidative, increase of GST activity, and low cell viability of
339 the roots.

340 From the data obtained, *E. densa*, *C. demersum*, and *M. aquaticum* were efficient in the
341 removal of the contaminants in the water and this system can be a successful tool to use in water
342 supply reservoirs. Two experimental large-scale Green Liver System® were built in China and
343 in Northeast of Brazil. Both experiments showed an excellent performance with an uptake
344 higher than 80 % for cyanotoxins and an antibiotic (Pflugmacher et al. 2015). According to the
345 current situation of Iraí Reservoir, the application of the Green Liver System® could be an
346 alternative to solve the problem of contamination, reduce costs for the water treatment, and
347 reduce the risk to human health.

348 **5. Conclusion**

349 Paracetamol, diclofenac, and MC-LR were quantified in the Iraí Reservoir and these
350 compounds can pose a risk to environmental and human health. In the present study, the
351 compounds were taken up by the plants tested and no oxidative stress incitement was evident
352 after 14 days. Therefore, *E. densa*, *C. demersum*, and *M. aquaticum* are deemed as suitable tools
353 to use in phytoremediation. In addition, the Green Liver System® is a sustainable method that
354 could be applied to improve the drink water treatment.

355 **6. Acknowledgments**

356 We thank Ms. S. Kühn (TU Berlin) and Ms. M. Behmanesh (TU Berlin) for assistance
357 in the laboratory. The research exchange (SLMC) to Germany was funded by CAPES
358 (Coordenação de Aperfeiçoamento de Pessoal de Nível Superior).

359

360 **Declaration of interest:**

361 Conflict of interest: none.

362

363 **6. References**

364 Amado LL, Monserrat JM. 2010. Oxidative stress generation by microcystin in aquatic
365 animals. Why and how. *Environ Int* 36(2):226-235.

366 <https://doi.org/10.1016/j.envint.2009.10.010>

367 Balsano E, Esterhuizen-Londt M, Hoque E, Pflugmacher S. 2015. Toxin resistance in aquatic
368 fungi poses environmentally friendly remediation possibilities: A study on the growth
369 responses and biosorption potential of *Mucor hiemalis* EH5 against cyanobacterial toxins.
370 *International Journal of Water and Waste Water Treatment* 1(1):1-9. doi [http://](http://dx.doi.org/10.16966/2381-5299.101)
371 dx.doi.org/10.16966/2381-5299.101

372 Baudhuin PH, Beaufay Y, Rahman-Li Y, Sellinger OH, Watliaux R, Jacques P, Duve C. 1964.
373 Tissue fractionation studies XVII. Intracellular distribution of monoamine oxidase,
374 aspartate amino-transferase, diaminoacid oxidase and catalase in rat liver tissue. *Biochem*
375 *J* 92(1):179-187.

376 Bittencourt-Oliveira MC. 2003. Detection of potential microcystin-producing cyanobacteria in
377 Brazilian reservoirs with a *mcyB* molecular marker. *Harmful Algae* 2(1):51-60.
378 [https://doi.org/10.1016/S1568-9889\(03\)00004-0](https://doi.org/10.1016/S1568-9889(03)00004-0)

379 Bittencourt-Oliveira MC, Hereman TC, Cordeiro-Araújo MK, Macedo-Silva I, Dias CT, Sasaki
380 FFC, Moura AN. 2013. Phytotoxicity associated to microcystins: a review. *Braz J Biol.*
381 74(4):753-760. <http://dx.doi.org/10.1590/1519-6984.06213>

382 Bouju E, Nastold P, Beck B, Hollender J, Corvini P. 2016. Elucidation of biotransformation of
383 diclofenac and 4'-hydroxydiclofenac during biological wastewater treatment. *J Hazard*
384 *Mater* 301:443-452. <https://doi.org/10.1016/j.jhazmat.2015.08.054>

385 Bradford MM. 1976. A rapid and sensitive method for the quantification of microgram
386 quantities of protein utilizing the principle of protein-dye binding. *Anal Biochem* 72:248-
387 254.

388 Brazil. Ministério da Saúde. Portaria n. 2.914 de 12 de dezembro de 2011. Diário Oficial da

389 União, n. 8239, seção 1, 2011.
390 http://site.sabesp.com.br/site/uploads/file/asabesp_doctos/PortariaMS291412122011.pdf

391 Calado SLM, Wojciechowski J, Santos GS, Magalhães VF, Padial AA, Cestari MM, Silva de
392 Assis H. 2017. Neurotoxins in a water supply reservoir: An alert to environmental and
393 human health. *Toxicol* 126:12-22. <https://doi.org/10.1016/j.toxicol.2016.12.002>

394 Contardo-Jara V, Kühn S, Pflugmacher S. 2015. Single and combined exposure to MC-LR
395 and BMAA confirm suitability of *Aegagropila linnaei* for use in Green Liver System® -
396 A case study with cyanobacterial toxins. *Aquat Toxicol* 165:101-108.
397 <http://dx.doi.org/10.1016/j.aquatox.2015.05.017>

398 Cunha SC, Pena A, Fernandes JO. 2017. Mussels as bioindicators of diclofenac contamination
399 in coastal environments. *Environ Pollut* 225:354-360.
400 <https://doi.org/10.1016/j.envpol.2017.02.061>

401 Ebele AJ, Abdallah MA, Hurrad S. 2017. Pharmaceuticals and personal care products (PPCPs)
402 in the freshwater aquatic environment. *Emerging Contaminants* 3:1-16.
403 <http://dx.doi.org/10.1016/j.emcon.2016.12.004>

404 Esterhuizen-Londt M, Schwartz K, Balsano E, Kühn S, Pflugmacher S. 2016. LC-MS/MS
405 method development for quantitative analysis of acetaminophen uptake by the aquatic
406 fungus *Mucor hiemalis*. *Ecotoxicol Environ Saf.* 128:230-235.
407 <https://doi.org/10.1016/j.ecoenv.2016.02.029>

408 Esterhuizen-Londt M, Hendel AL, Pflugmacher S. 2017. Mycoremediation of diclofenac using
409 *Mucor hiemalis*. *Environ Toxicol Chem* 99:798-808.
410 <https://doi.org/10.1080/02772248.2017.1296444>

411 European Commission. 2012. Revised directive of the European Parliament and of the Council
412 on priority substances in the field of water quality.
413 <http://ec.europa.eu/environment/water/water-danger>.

414 Fernández-Fuego D, Keunen E, Cuypers A, Bertrand A, González A. 2017. Mycorrhization

415 protects *Betula pubescens* Ehr. from metal-induced oxidative stress increasing its
416 tolerance to grow in an industrial polluted soil. J Hazard Mater 336:119-127.
417 <https://doi.org/10.1016/j.jhazmat.2017.04.065>

418 Fernandes LF, Lagos PED, Wosiack AC, Pacheco CV, Domingues L, Zenhder-Alves L,
419 Coquemala V. 2005. Comunidades fitoplanctônicas em ambientes lênticos. In: Andreoli
420 CV, Carneiro C, editors. Gestão integrada de mananciais de abastecimento eutrofizados.
421 Curitiba (Brazil): Sanepar-Finep. p. 500.

422 Ferrão-Filho AS, Herrera NA, Echeverri LF. 2014. Microcystin accumulation in cladocerans:
423 First evidence of MC uptake from aqueous extracts of a natural bloom sample. Toxicon
424 87:26–31. <https://doi.org/10.1016/j.ecoenv.2008.02.002>

425 Flores-Rojas NC, Esterhuizen-Londt M, Pflugmacher S. 2015. Antioxidative stress responses
426 in the floating macrophyte *Lemna minor* L. with cylindrospermopsin exposure. Aquat
427 Toxicol 169:188-195. <https://doi.org/10.1016/j.aquatox.2015.11.002>

428 Fonte P, Reis S, Sarmiento B. 2016. Facts and evidences on the lyophilization of polymeric
429 nanoparticles for drug delivery. J Control Release. 225:75-86.
430 <https://doi.org/10.1016/j.jconrel.2016.01.034>

431 Gibble CM, Peacock MB, Kudela RM. 2016. Evidence of freshwater algal toxins in marine
432 shellfish: Implications for human and aquatic health. Harmful Algae 59:59-66.
433 <https://doi.org/10.1016/j.hal.2016.09.007>

434 Gill SS, Tuteja N. 2010. Reactive oxygen species and antioxidant machinery in abiotic stress
435 tolerance in crop plants. Plant Physiol Biochem. 48:909-930.
436 <https://doi.org/10.1016/j.plaphy.2010.08.016>

437 Gröner F, Höhne C, Kleiner W, Kloas W. 2017. Chronic diclofenac exposure affects gill
438 integrity and pituitary gene expression and displays estrogenic activity in Nile Tilapia
439 (*Oreochromis niloticus*). Chemosphere 166:473-481.
440 <https://doi.org/10.1016/j.chemosphere.2016.09.116>

441 Guiloski IC, Ribas JL, Piancini LDS, Dagostim AC, Calado SLM, Fávaro LF, Boschen SL,
442 Cestari MM, Cunha C, Silva de Assis HC. 2017a. Effects of environmentally relevant
443 concentrations of the anti-inflammatory drug diclofenac in freshwater fish *Rhamdia*
444 *quelen*. *Ecotoxicol Environ Saf.* 139:291-300.
445 <https://doi.org/10.1016/j.ecoenv.2017.01.053>

446 Guiloski IC, Ribas JL, Piancini LDS, Dagostim AC, Cirio SM, Fávaro LF, Boschen SL, Cestari
447 MM, Cunha C, Silva de Assis HC. 2017b. Paracetamol causes endocrine disruption and
448 hepatotoxicity in male fish *Rhamdia quelen* after subchronic exposure. *Environ Toxicol*
449 *Pharmacol.* 53:111-120. <https://doi.org/10.1016/j.etap.2017.05.005>

450 Gupta N, Pant SC, Vijayraghavan R, Rao PVL. 2003. Comparative toxicity evaluation of
451 cyanobacterial cyclic peptide toxin microcystin variants (LR, RR, YR) in mice.
452 *Toxicology* 188:285–296. [https://doi.org/10.1016/S0300-483X\(03\)00112-4](https://doi.org/10.1016/S0300-483X(03)00112-4)

453 Habig W, Pabst MJ, Jacoby WB. 1974. Glutathione S-transferase: the first step in mercapturic
454 acid formation. *J Biol Chem.* 249:1730-1739.

455 Hauser-Davis RA, Lavradas RT, Lavandier RC, Rojas EGAR, Guarino AWS, Ziolli RL. 2015.
456 Accumulation and toxic effects of microcystin in tilapia (*Oreochromis niloticus*) from an
457 eutrophic Brazilian lagoon. *Ecotoxicol Environ Saf.* 112:132-136.
458 <https://doi.org/10.1016/j.ecoenv.2014.10.036>

459 Huber C, Bartha B, Harpaintner P, Schroder P. 2009. Metabolism of acetaminophen
460 (paracetamol) in plants – two independent pathways result in the formation of a
461 glutathione and a glucose conjugate. *Environ Sci Pollut Res* 16:206-213.
462 <http://doi.org/10.1007/s11356-008-0095-z>

463 Huber C, Bartha B, Schroder P. 2012. Metabolism of diclofenac in plants – hydroxylation is
464 followed by glucose conjugation. *J Hazard Mater* 243:250-256.
465 <https://doi.org/10.1016/j.jhazmat.2012.10.023>

466 Jewell KS, Falás P, Wick A, Joss A, Temes TA. 2016. Transformation of diclofenac in hybrid
467 biofilm-activated sludge processes. *Water Res* 105:559-567.
468 <https://doi.org/10.1016/j.watres.2016.08.002>

469 Kramer RD, Mizukawa A, Ide AH, Marcante LO, Dos Santos MM, De Azevedo JCR. 2015.
470 Determinação de anti-inflamatórios na água e sedimento e suas relações com a qualidade
471 da água na bacia do Alto Iguaçu, Curitiba-PR. *Rev. bras. epidemiol* 20:667-667.

472 Kummerová M, Zezulka S, Babula P, Tríska J. 2016. Possible ecological risk of two
473 pharmaceuticals diclofenac and paracetamol demonstrated on a model plant *Lemna*. *J*
474 *Hazard Mater* 302:351-361. <http://dx.doi.org/10.1016/j.jhazmat.2015.09.057>

475 Lajayer BA, Ghorbanpour M, Nikabadi S. 2017. Heavy metals in contaminated environment:
476 Destiny of secondary metabolite biosynthesis, oxidative status and phytoextraction in
477 medical plants. *Ecotoxicol Environ Saf.* 145:377-390.
478 <https://doi.org/10/1016/j.ecoenv.2017.07.035>

479 Liu J, Sun Y. 2015. The role of PP2A-associated proteins and signal pathways in microcystin-
480 LR toxicity. *Toxicol Lett* 236:1-7. <http://dx.doi.org/10.1016/j.toxlet.2015.04.010>

481 Liu Y, Wang L, Pan B, Wang C, Bao S, Nie X. 2017. Toxic effects of diclofenac on life history
482 parameters and the expression of detoxification-related genes in *Daphnia magna*. *Aquat*
483 *Toxicol* 183:104-113. <https://doi.org/10.1016/j.aquatox.2016.12.020>

484 Lonappan L, Brar SK, Das RK, Verma M, Surampalli RY. 2016. Diclofenac and its
485 transformation products: Environmental occurrence and toxicity - A review. *Environ Int*
486 96:127-138. <https://doi.org/10.1016/j.envint.2016.09.014>

487 Matamoros V, Nguyen X, Arias CA, Salvadó V, Brix H. 2012. Evaluation of aquatic plants for
488 removing polar microcontaminants: A microcosm experiment. *Chemosphere* 88:1257-
489 1264. <https://doi.org/10.1016/j.chemosphere.2012.04.004>

490 Näslund J, Fick J, Asker N, Ekman E, Larsson J, Norrgren L. 2017. Diclofenac affects kidney
491 histology in the three-spined stickleback (*Gasterosteus aculeatus*) at low µg/L
492 concentrations. *Aquat Toxicol* 186:87-96. <https://doi.org/10.1016/j.aquatox.2017.05.017>

493 Nimptsch J, Wiegand C, Pflugmacher S. 2008. Cyanobacterial toxin elimination via
494 bioaccumulation of MCLR in aquatic macrophytes: An application of the “Green Liver
495 Concept”. *Environ Sci Technol* 42:8552-8557. <http://dx.doi.org/10.1021/es8010404>

496 Nunes B, Antunes S, Santos J, Martins L, Castro BB. 2014. Toxic potential of paracetamol to
497 freshwater organisms: A headache to environmental regulators? *Ecotoxicol Environ Saf.*
498 107:178-185. <https://doi.org/10.1016/j.ecoenv.2014.05.027>

499 Nunes B, Nunes J, Soares AMVM, Figueira E, Freitas R. 2017. Toxicological effects of
500 paracetamol on the clam *Ruditapes philippinarum*: exposure vs recovery. *Aquat Toxicol*
501 192:198-206. <http://dx.doi.org/10.1016/j.aquatox.2017.09.015>

502 Omid A, Esterhuizen-Londt M, Pflugmacher S. 2018. Still challenging: the ecological function
503 of the cyanobacterial toxin microcystin – What we know so far. *Toxin Reviews* 37(2):
504 87-105.

505 Pflugmacher S. 2004. Promotion of oxidative stress in *C. demersum* due to exposure to
506 cyanobacterial toxin. *Aquat Toxicol* 3:169-178.
507 <https://doi.org/10.1016/j.aquatox.2004.06.010>

508 Pflugmacher S, Kühn S, Lee S, Choi J, Baik S, Kwon K, Contardo-Jara V. 2015. Green Liver
509 Systems® for water purification: Using the phytoremediation potential of aquatic
510 macrophytes for the removal of different cyanobacterial toxins from water. *Am J Plant*
511 *Sci* 6:1607-1618. <http://dx.doi.org/10.4236/ajps.2015.69161>

512 Pflugmacher S, Kwon KS, Baik S, Kim S, Kühn S, Esterhuizen-Londt M. 2016. Physiological
513 responses of *Cladophora glomerata* to cyanotoxins: a potential new phytoremediation
514 species for the Green Liver Systems. *Toxicol Environ Chem* 98:241-259.
515 <https://doi.org/10.1080/02772248.2015.1119835>

516 Pflugmacher S, Olin M, Kankaanpää H. 2007. Nodularin induces oxidative stress in the Baltic
517 Sea brown alga *Fucus vesiculosus* (Phaeophyceae). *Mar Environ Res.* 64:149–159.
518 <https://doi.org/10.1016/j.marenvres.2006.12.011>

519 Pitard FF. 1993. Pierre Gy's Sampling theory and sampling practice. CRC Press LLC, Boca
520 Raton, Florida.

521 Pradhan A, Silva CO, Silva C, Pascoal C, Cássio F. 2016. Enzymatic biomarkers can portray
522 nanoCuO-induced oxidative and neuronal stress in freshwater shredders. *Aquat Toxicol*
523 180:227-235. <https://doi.org/10.1016/j.aquatox.2016.09.017>

524 Ranieri E, Verlicchi P, Young TM. 2011. Paracetamol removal in subsurface flow constructed
525 wetlands." *J Hydrol* 404:130-135. <https://doi.org/10.1016/j.jhydrol.2011.03.015>

526 Reis AR, Tabei K, Sakakibara Y. 2014. Oxidation mechanism and overall removal rates of
527 endocrine disrupting chemicals by aquatic plants. *J Hazard Mater* 265:79-88.
528 <http://dx.doi.org/10.1016/j.jhazmat.2013.11.042>

529 Romero-Oliva CS, Contardo-Jara V, Pflugmacher S. 2015. Time dependent uptake,
530 bioaccumulation and biotransformation of cell free crude extract microcystins from Lake
531 Amatitlán, Guatemala by *Ceratophyllum demersum*, *Egeria densa* and *Hydrilla*
532 *verticillata*. *Toxicon*, 105:62-73, [10.1016/j.toxicon.2015.08.017](https://doi.org/10.1016/j.toxicon.2015.08.017)

533 Santos LHLM, Araújo AN, Fachini A, Pena A, Delerue-Matos C, Montenegro MCBSM. 2010.
534 Ecotoxicological aspects related to the presence of pharmaceuticals in the aquatic
535 environment. *J Hazard Mater* 175:45-95. <https://doi.org/10.1016/j.jhazmat.2009.10.100>

536 Schulz R, Bundschuh M, Gergs R, Bruehl CA, Entling MH, Fahse L, Fror O, Jungkunst HF,
537 Lorke A, Schafer RB, Schaumann GE, Schwenk K. 2015. Review on environmental
538 alterations propagating from aquatic to terrestrial ecosystems. *Sci Total Environ* 538:246-
539 261. <https://doi.org/10.1016/j.scitotenv.2015.08.038>

540 Scholz SN, Esterhuizen-Londt M, Pflugmacher S. 2017. Rise of toxic cyanobacterial blooms in
541 temperate freshwater lakes: causes, correlations and possible countermeasures. *Environ*

542 Toxicol Chem 99:543-577.

543 Spengler A, Wanninger L, Pflugmacher S. 2017. Oxidative stress mediated toxicity of TiO₂
544 nanoparticles after a concentration and time dependent exposure of the aquatic
545 macrophyte *Hydrilla verticillata*. *Aquat Toxicol* 190:32-39.
546 <https://doi.org/10.1016/j.aquatox.2017.06.006>

547 Van Der Oost R, Beyer J, Vermeulen NPE. 2003. Fish bioaccumulation and biomarkers in
548 environmental risk assessment: a review. *Environ Toxicol Pharmacol* 13:57-149.
549 [https://doi.org/10.1016/S1382-6689\(02\)00126-6](https://doi.org/10.1016/S1382-6689(02)00126-6)

550 Vilvert E, Contardo-Jara V, Esterhuizen-Londt M, Pflugmacher S. 2017. The effect of
551 oxytetracycline on physiological and enzymatic defense responses in aquatic plant
552 species *Egeria densa*, *Azolla caroliniana*, and *Taxiphyllum Barbieri*. *Toxicol Environ*
553 *Chem* 99:104-116. <https://doi.org/10.1080/02772248.2016.1165817>

554 WHO Guidelines for Drinking-Water Quality. 1998. Health Criteria and Other Supporting
555 Information. Addendum, World Health Organization, Geneva. Second Edition, Vol. 2.
556 http://www.who.int/water_sanitation_health/dwq/2edaddvol2a.pdf

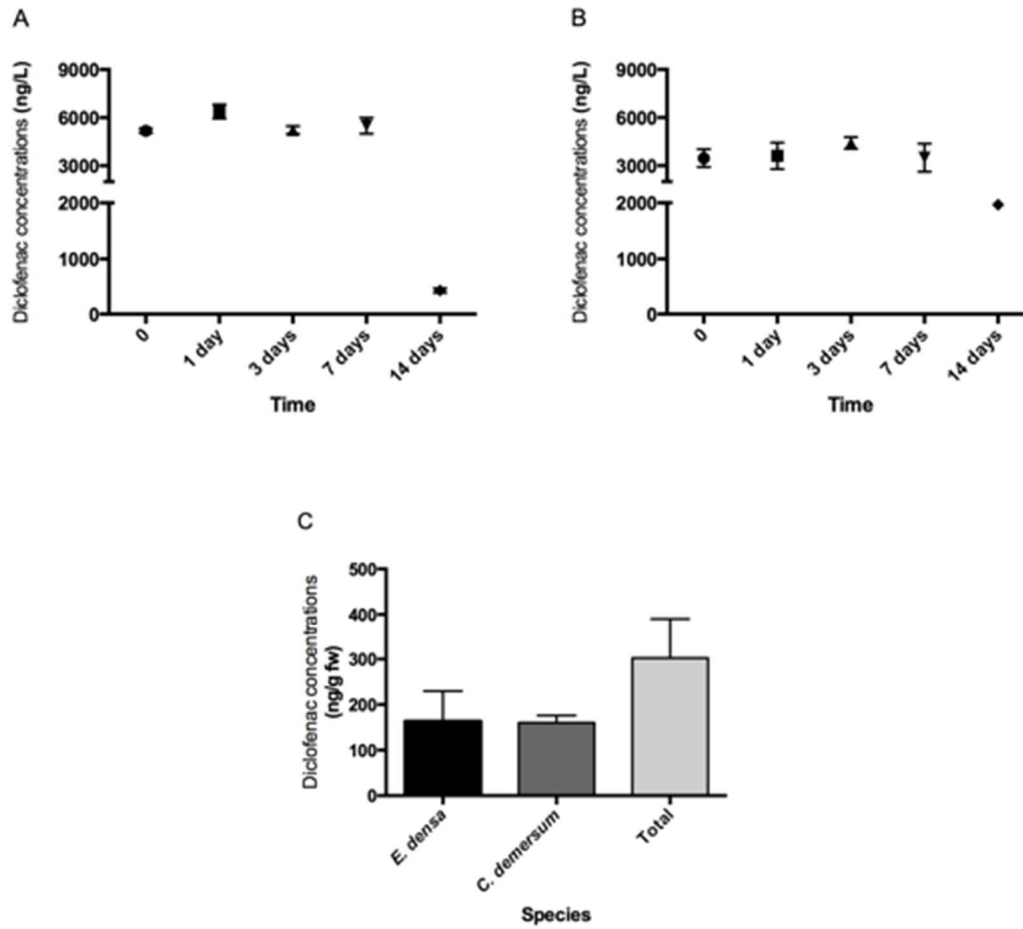
557 Yang Y, Ok YS, Kim K, Kwon EE, Tsang YF. 2017. Occurrences and removal of
558 pharmaceuticals and personal care products (PPCPs) in drinking water and water/sewage
559 treatment plants: A review. *Sci Tot Environ* 596-597:303-320.
560 <https://doi.org/10.1016/j.scitotenv.2017.04.102>

561 Yuan M, Carmichael WW, Hilborn ED. 2006. Microcystin analysis in human sera and liver
562 from human fatalities in Caruaru, Brazil 1996. *Toxicol* 48:627-640.
563 <https://doi.org/10.1016/j.toxicol.2006.07.031>

564 Zegura B, Straser A, Filipic M. 2011. Genotoxicity and potential carcinogenicity of
565 cyanobacterial toxins – a review. *Mutat Res* 727:16-41. doi:10.1016/j.mrrev.2011.01.002

566

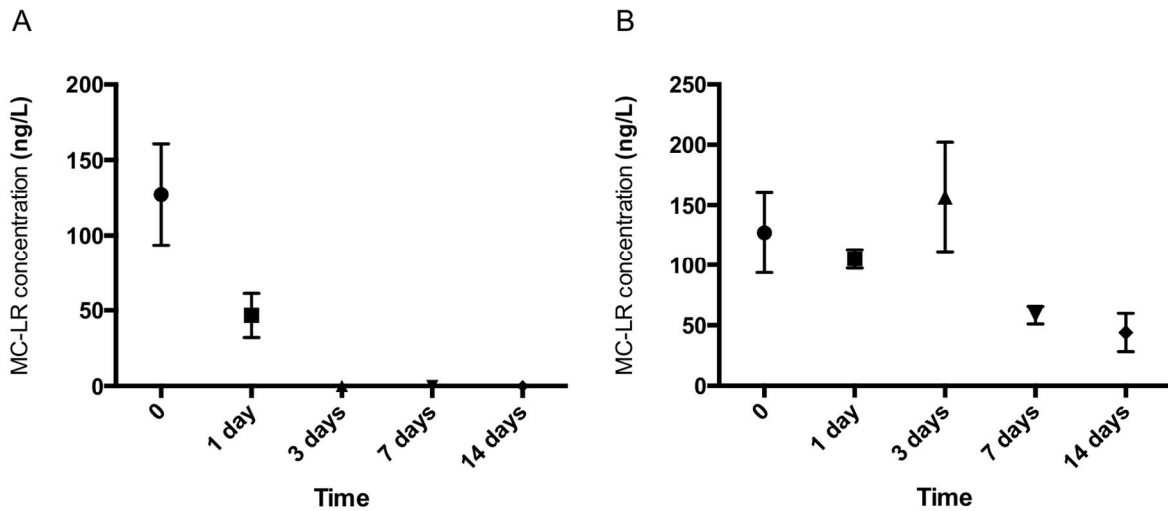
567 **Figure captions**



568

569 **Figure 1.** Diclofenac concentration: (A) Experiment Green Liver System®; (B) Experiment
570 control without plants and (C) plant species from experiment Green Liver System®. Mean
571 \pm SD (n=5).

572

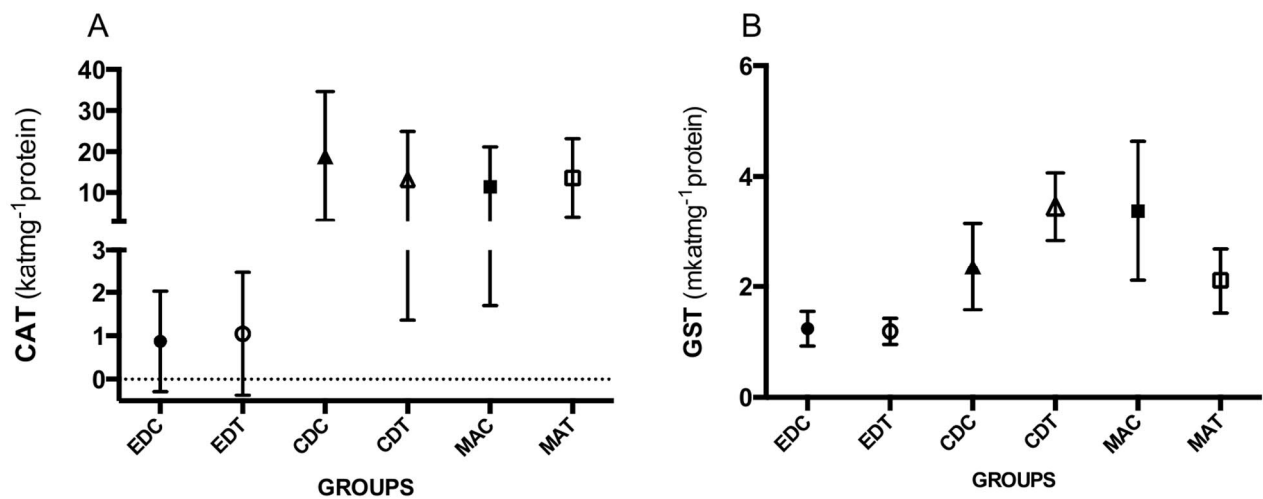


573

574

575

Figure 2. MC-LR concentrations: (A) Experiment Green Liver System®; (B) Experiment control without plants. Mean \pm SD (n=5).



576

577

578

579

580

Figure 3. CAT activity (A) and GST activity (B) in plant tissues with exposure. Data points represent mean \pm SD (n=3); $p < 0.05$. **EDC:** *E. densa* control; **EDT:** *E. densa* treatment; **CDC:** *C. demersum* control; **CDT:** *C. demersum* treatment; **MAC:** *M. aquaticum* control; **MAT:** *M. aquaticum* treatment.