1	This manuscript is contextually identical with the following published paper:					
2	Bruna PaolinelliReis ^{a1} , Sebastião Venâncio Martins ^a , Elpídio Inácio Fernandes Filho ^b ,					
3	Tathiane Santi Sarcinelli ^c , José MarinaldoGleriani ^a , Helio Garcia Leite ^a , Melinda					
4	Halassy ^d 2019. Forest restoration monitoring through digital processing of high					
5	resolution images. Ecological Engineering Volume 127, February 2019, Pages 178-186.					
6	The original published pdf available in this website:					
7	https://doi.org/10.1016/j.ecoleng.2018.11.022					
8						
9						
10 11	FOREST RESTORATION MONITORING THROUGH DIGITAL PROCESSING OF HIGH RESOLUTION IMAGES					
12						
13 14 15	Bruna Paolinelli Reis ^{a1*} ; Sebastião Venâncio Martins ^a ; Elpídio Inácio Fernandes Filho ^b ; Tathiane Santi Sarcinelli ^c ; José Marinaldo Gleriani ^a ; Helio Garcia Leite ^a ; Melinda Halassy ^d					
17 18 19 20 21 22 23 24 25 26 27	 ^a Forest Engineer Department. Universidade Federal de Viçosa. Avenida P. H. Rolfs s/n, CEP 36570-000, Viçosa, MG, Brasil. ^b Soil Department. Universidade Federal de Viçosa. Avenida P. H. Rolfs s/n, CEP 36570-000, Viçosa, MG, Brasil. ^c Fibria Celulose S.A. Meio Ambiente Florestal. Rodovia Aracruz - Barra do Riacho, s/n, km 25, Zipcode: 29.197-900. Aracruz, ES, Brasil. ^d MTA Centre for Ecological Research, Institute of Ecology and Botany, Alkomány u. 2-4, 2163, Vácrátót, Hungary *corresponding author. <i>E</i> mail advass: brunapaolinelli@amail.com (P. P. Pais). 					
28 29	Abstract					
30	Monitoring and evaluating forest restoration projects is a challenge especially in large-					

32 and multitemporal data allows us to gauge the restoration success with more accurately

¹Permanet address: Eötvös Loránd University, Department of Plant Taxonomy, Ecology and Theoretical Biology, Pázmány P. stny. 1/C, 1117 Budapest, Hungary

33 and in small time. The objective of this study was to elaborate and compare methods of 34 remote monitoring of forest restoration using Light Detection and Ranging (LIDAR) 35 data and multispectral imaging from Unmanned Aerial Vehicle (UAV) camera, in 36 addition to comparing the efficiency of supervised classification algorithms Maximum 37 Likelihood (ML) and Random Forest (RF). The study was carried out in a restoration 38 area with about 74 hectares and five years of implementation, owned by Fibria Celulose 39 S.A., in the southern region of Bahia State, Brazil. We used images from Canon S110 40 NIR (green, red, Near Infrared) on UAV and LIDAR data composition (intensity image, 41 Digital Surface Model, Digital Terrain Model, normalized Digital Surface Model). The 42 monitored restoration indicator was the land cover separated in three classes: canopy 43 cover, bare soil and grass cover. The images were classified using the ML and RF 44 algorithms. To evaluate the accuracy of the classifications, the Overall Accuracy (OA) 45 and the Kappa index were used, and the last was compared by Z test. The area occupied 46 by different land cover classes was calculated using ArcGIS and R. The results of OA, 47 Kappa and visual evaluation of the images were excellent in all combinations of the 48 imaging methods and algorithms analyzed. When Kappa values for the two algorithms 49 were compared, RF presented better performance than ML with significant difference, 50 but when sensors (UAV camera and LIDAR) were compared, there were no significant 51 differences. There was little difference between the area occupied by each land cover 52 classes generated by UAV and LIDAR images. The highest cover was generated for 53 canopy cover followed by grass cover and bare soil in all classified images, indicating 54 the need of adaptive management interventions to correct the area trajectory towards the 55 restoration success. The methods employed in this study are efficient to monitor 56 restoration areas, especially on a large scale, allowing us to save time, fieldwork and 57 invested resources.

58

59 Keywords: Light Detection and Ranging (LIDAR), Unmanned Aerial Vehicle (UAV),

60 Random Forest (RF), Maximum Likelihood Algorithm, Recovery of Degraded Areas,

- 61 forest restoration
- 62

63 1. Introduction

Due to the high demand for environmental regularization of big companies and farmers and the need to mitigate environmental impacts generated by human activities, restoration projects have been increased across the globe (Li et al., 2017). The expansion of works and techniques used in restoration initiatives and successive evaluations of what was done in the past made it possible to correct and to adapt the previously used methods to favor functional ecosystem reestablishment (Rodrigues et al., 2009).

71 Long-term follow-up of large-scale restoration projects is critical because it 72 allows us to evaluate their success or to correct their trajectory through the generation of 73 ecological management recommendations (enrichment, invasive species control, and 74 othres). Therefore, the application of these corrective actions can guarantee the success 75 of the project already in the first years of the planting (Ruiz-Jaen and Aide, 2005; Melo 76 et al., 2013; Zahawi et al., 2015). However, monitoring restoration is a challenge 77 especially in large-scale projects (Viani et al., 2017, Ockendon et al, 2018), mainly 78 because the field assessment of several indicators can be time consuming, costly, and 79 require trained technicians (Zahawi et al., 2015; Reif and Theel, 2017). Additionally, a 80 clear definition of which indicator should be measured and the frequency of assessment 81 still lack (Rodrigues et al., 2009). Alternative approaches to evaluating and monitoring 82 different restoration parameters using remote sensing and digital image processing 83 techniques are considered promising to reduce the need for field measurement methods 84 (Reif and Theel, 2017; Viani et al., 2017), especially for large scale restoration projects 85 and for areas that are difficult to access (Mascaro et al., 2014; Zahawi et al., 2015). 86 Additionally, with the use of remote monitoring it is possible to evaluate restoration in 87 landscape scale, not only in selected samples, as it is usually done in field monitoring 88 (Zahawi et al., 2015).

Land cover composition is an important indicator to evaluate landscape
condition and to monitor status and trends of ecosystem change over a specific time
(Xian et al., 2009). In our study, land cover indicator was separated in three classes:
canopy cover, grass cover and bare soil. According to Harris et al. (2012), canopy cover

93 influences of light entering forest structure, amount the forest, water 94 infiltration, facilitates the growth of understory native seedlings and controls of 95 undesirable grasses, among others. The evaluation of invasive grass cover is considered 96 of high importance in the early stages of the forest restoration process, because these 97 species are spreading rapidly and compete with native species impeding or hindering 98 their growth (Rocha-Nicoleite et al, 2017). Bare soil should be monitored because when 99 it occurs in large spots also compromises the seedlings survival preventing the water 100 infiltration and inducing soil erosion and nutrient loss, among others (Muñoz-Rojas et 101 al., 2016).

102 Out of the new remote sensing techniques that have been widespread Light 103 Detection and Ranging (LIDAR) enables to determine the distance between the sensor 104 and the target surface using laser pulses (Lefsky et al., 2002). Additionally, this sensor 105 is able to generate the pulse return intensity image, which can be useful to classify land 106 cover (Song et al., 2002). According to Giongo et al. (2010) this sensor has proved to be 107 efficient for different forestry applications, since it allows the measurement of canopy 108 topography (Dubayah and Drake, 2000), biomass (Bortolot and Wynne, 2005), volume 109 (Ioki et al., 2010), species identification (Kim et al., 2007; Holmgren et al., 2008) and 110 several other applications. Another remote sensing technique that has been widely used 111 is high-resolution imagery acquired from Unmanned Aerial Vehicle (UAV) camera. 112 Several studies have demonstrated the efficiency of these image processing, e.g. in the 113 mapping of invasive species (Michez et al., 2016), land use and cover planning (Silva 114 et. al., 2016), to monitor tropical forest recovery (Zahawi et al., 2015) and biodiversity 115 (Koh and Wich, 2012; Getzin et al. 2012; Paneque-Galves et al., 2014).

Supervised classification is a process of image information extraction where theuser selects representative samples of different classes found in the image (Campbell,

118 1996). A widely used algorithm for image classification is the Maximum Likelihood 119 (ML), which considers the classes involved in a Gaussian probability density function 120 (Hagner and Reese, 2007). Another more robust algorithm that has been shown to be 121 very efficient in the classification of satellite images is the Random Forest (RF) 122 (Gislason et al., 2006; Lawrence et al., 2006; Puissant et al., 2014). This algorithm is 123 based on the generation of multiple decision trees that vote for the most popular class 124 (Breiman, 2001) to produce more accurate classification (Cutler et al., 2007; Puissant, 125 2014).

The aim of the present study was 1) to elaborate and compare methods of remote monitoring of forest restoration using images from LIDAR data and UAV camera, 2) to compare the efficiency of supervised classification algorithms ML and RF and 3) to obtain land cover composition by classified images.

130

131

132

133 2. Materials and methods

134 *2.1. Study site*

The present study was carried out in conservation areas under restoration process
owned by Fibria Celulose S.A., on a farm named "Project Maria Mirreis T734", 39.68°
S, 17.71° W, located in the municipality of Caravelas, extreme south of Bahia state,
Brazil (Figure 1).



139

Fig. 1. Location and Land Use of the study site called "Project T734 Maria Mirreis", in
south of Bahia state, Brazil.

According to the classification of Köppen, the region is in the transition between Tropical rainforest climate, hot and humid tropical climate on the coast and seazonal climate, with dry winter in the interior (Zonete et al., 2010). The average annual rainfall is around 1,100 mm (Almeida et al., 2008) and the average temperature is 25°C, without defined dry season (Costa et al., 2009). The predominant soils in the region are the kaolinitic Latosols and Yellow or Red-Yellow Argisols, presenting low fertility, and, in case of Argisols, presence of a dense layer in the subsurface (Moreau et al., 2006). The
natural vegetation in the region includes Atlantic forests (Dense Ombrophilous Forest
and Seasonal Semideciduous Lowland Forests) (Saporetti, 2012).

152 The farm that included our study area has a total area of 664.42 ha, where 432.09 ha are planted with eucalyptus, 207.17 ha is conservation areas, 15.78 ha of 153 154 roads and 9.38 ha for various uses. Among the conservation areas, there are natural 155 forests in different successional stages and degraded pastures dominated by the exotic 156 grass Urochloa decumbens (Stapf) R.D.Webste (74 ha). Restoration projects are being 157 carried out mainly in these degraded pastures, which were originally occupied by Dense 158 Ombrofilous forest, and native grasses do not occur (Veloso, 1991; Ivanauskas and 159 Assis, 2012). Aiming to achieve the restoration success, interventions were 160 implemented in 2010 through the methodologies of natural regeneration or planting 161 native species seedling.

162 *2.2. Image gathering from UAV and LIDAR*

Aerial images of the study area were obtained from Canon S110 NIR cameras, coupled to UAV in the spring season. The resolution of the images captured by the cameras is 4000 x 3000 pixels, and the spatial resolution is 0.08 meters. The camera captures images in the spectral region of green, red, and near infrared (NIR). The Normalized Difference Vegetation Index (NDVI) (Rouse et al, 1973) (Equation 1) and Soil Adjusted Vegetation Index (SAVI) (Huete, 1988) (Equation 2) were calculated. Then, three spectral bands and two spectral transformations were combined.

170

171
$$NDVI = \frac{(NIR-Red)}{(NIR+Red)}$$
 (Equation 1)

172 where: *NIR* = Near Infrared band and Red = Red band

174

75
$$SAVI = \frac{(1+Ls)*(NIR-Red)}{(Ls+NIR+Red)}$$
 (Equation 2)

176 where: Ls is a constant denominated factor of adjustment of the SAVI index, being able 177 to assume values from 0.25 to 1 depending on the land cover. In this case, the value of 178 0.5 was considered, representing vegetation with intermediate density (Huete, 1988)

179

180 Images were also obtained through a RIEGL LMS - Q680I Airborne Laser 181 system denominated LIDAR, which works on Near Infrared region with a wavelength 182 of 1,036 nm. The images were taken in spring season. From LIDAR point clouds, it was 183 possible to generate the pulse return intensity image, Digital Surface Model (DSM), 184 Digital Terrain Model (DTM) and normalized Digital Surface Model (nDSM) using 185 ArcGIS 10.2 (ESRI, 2015). nDSM stores information of the objects height, obtained 186 through the difference between DSM and DTM. According to Song et al. (2002), the 187 intensity is defined as the ratio of strength of reflected light to the emitted light and can 188 be useful to classify the land cover. By using the obtained data, a 0.5 m spatial 189 resolution image composed of four bands (Intensity, nDSM, DTM, DSM) was 190 generated. Thus, in addition to the information of the vegetation height, there is also the 191 spectral information related to the pulse return intensity.

192 2.3.Land cover monitoring

193 The monitored restoration indicator was the land cover separated in three 194 classes: canopy cover, bare soil and grass cover. These classes were chosen due to the 195 feasibility for obtaining results from aerial images classification. All these classes are 196 easily identified by visual image interpretation and digital image processing due to clear 197 differences between their spectral signatures. They also have ecological importance in 198 early stages of restoration process and relevance for generating adaptive management

recommendations, which aim to correct the trajectory of a restoration area allowing itsecological succession (Atlantic Forest Restoration Pact, 2013; Viani et al., 2017).

UAV camera is a passive sensor, which requires sunlight as an illumination source for obtaining images. Therefore, the presence of clouds and shadows modifies the reflectance of the objects and can reduce the classification accuracy. For this reason, besides the three mentioned classes, it was necessary to establish another class representing objects in shadows, which was called "shadow".

206

207 2.4. Sample selection, classification, accuracy, comparison and class area

208 *quantification*

209 Fifty representative samples of each class (canopy cover, bare soil and grass 210 cover, shadow) were selected over UAV camera images. The same number of samples 211 per class (canopy cover, bare soil and grass cover) were chosen over LIDAR return 212 intensity channel. These samples were polygons with the same size, 450 pixels on 213 average for the UAV camera images and 55 pixels for LIDAR images due to the 214 different spatial resolution between both sensors. The software ArcGIS 10.2 (ESRI, 215 2015) was used for sample acquisition. Seventy percent of the samples polygons were 216 randomly selected for training and thirty percent them was left for validation of both 217 algorithms (ML and RF).

The supervised classification using ML algorithm was done with ArcGIS 10.2 software, whereas the RF algorithm was done with R 3.3.2 software (R Core Team, 2016). The fifty samples used for classification were randomly divided into training and validation samples ten times, and for each division classification models were fitted. RF algorithm also permitted to generate the variable importance, which represents the most

important variables for creating algorithm's decision trees, through two methods:randomization and Gini Index, described by Hastie et al. (2009).

225 Thirty percent of the selected samples from UAV and LIDAR images were used 226 for the validation phase to verify the accuracy of the methods. The classification 227 accuracy was assessed with the confusion matrix approach. From this matrix, 228 classification accuracy indicators, such as Producer's accuracy, User's accuracy, 229 Overall classification accuracy, and Kappa coefficient, were calculated. Additionally, 230 ten samples of each class were collected in the field, totaling 30 validation samples, and 231 the coordinates of each sampling point was obtained with GPS. This sampling was done 232 to evaluate whether the sampling on UAV and LIDAR images may have some bias that 233 leads to over or underestimated accuracy. After obtaining the Kappa Index, it was 234 possible to classify its value according to the literature (Landis and Kock, 1977).

As it was mentioned before 10 classifications were performed by each combination between imaging method and algorithm, therefore we calculate the average of Producer's accuracy, User's accuracy, Overall classification accuracy, and Kappa coefficient, in each situation. The Z test was used to compare Kappa values among all different classified images with a 95% confidence level, Z> 1.96 (Foody, 2004; Congalton and Green, 2009).

We calculated the land cover composition based on percentage class area (canopy cover, grass cover, bare soil and shadow) generated by the most accurate classification (from the 10 classifications), for each classified image using ArcGIS software (ESRI, 2015) and R software (R Core Team, 2016) for ML and RF, respectively. The complete procedure that we used to classify UAV and LIDAR images with different classifying algorithms and the validation performed to verify the classifiers accuracy is presented in Figure 2.



Fig 2. Flowchart representing the methodology used to classify LIDAR and UAV
images, using the ML and RF algorithms, and the accuracy of the classifications. DTM:
Digital Terrain Model, DSM: Digital Surface Model, LIDAR: Light Detection and
Ranging, ML: Maximum Likelihood, nDSM: normalized Digital Surface Model, NDVI:
Normalized Difference Vegetation Index, RF: Random Forest, SAVI: Soil Adjusted
Vegetation Index, UAV: Unmanned Aerial Vehicle.

- 257 **3.** Results
- 258 3.1. Classification by the Maximum Likelihood and Random Forest algorithms from
- 259 *UAV camera images*
- 260 The images obtained by UAV camera classified using ML and RF algorithms is
- 261 presented in Figure 3, and it represents part of the restoration project.



262

Fig. 3. a) UAV camera images (Red, Green, Near Infrared, NDVI, SAVI), which
represent part of the study area; b) Image classified by Maximum Likelihood algorithm;
c) Image classified by Random Forest algorithm.

- 266
- 267 It is noticed that all the classes were well delimited on UAV classified images by
- both algorithms and they were quite similar (Fig.3).
- The Tables 1 and 2 represents the mean values generated from the ten confusion
- 270 matrices for both classifiers. It was observed that Kappa and Overall Accuracy were

- 271 high and ranked as excellent (Landis and Kock, 1977) for both algorithms. The classes
- 272 separation was also assessed through user's and producer's accuracy. The values of
- these indices were high in both classifications, indicating a good level of discrimination
- between classes. Additionally, "shadow" was the class that most generates confusion,
- especially with "bare soil" and "grass cover" for both classifiers.

276 Table 1

277 Mean values generated from the ten confusion matrices, with the Maximum Likelihood278 classifier, for UAV camera images

CLASS	Bare Soil	Canopy cover	Grass cover	Shadow	Total	PA (%)
Bare soil	6463	3	91	388	6945	93.17
Canopy cover	0	8097	184	94	8375	96.75
Grass cover	21	65	7929	440	8455	94.02
Shadow	376	282	114	3858	4630	81.55
Total	6860	8447	8318	4779	28404	
UA (%)	96.21	96.45	95.63	90.00		
Average Kappa					0.90	
Average OA					0.93	

279 OA: Overall Accuracy, PA: Producer's Accuracy, UA: User's Accuracy.

280

281 Table 2

282 Mean values generated from the ten confusion matrices, with the Random Forest283 classifier, for UAV camera images

CLASS	Bare Soil	Canopy cover	Grass cover	Shadow	Total	PA (%)
Bare soil	6408	3	43	0	6454	99.34
Canopy cover	1	6992	109	1	7103	98.65
Grass cover	186	244	5836	172	6438	89.43
Shadow	1	80	63	2493	2637	93.88
Total	6596	7319	6051	2666	22632	
UA (%)	96.46	95.35	96.84	93.12		
Average Kappa					0.94	
Average OA					0.96	

²⁸⁴ OA: Overall Accuracy, PA: Producer's Accuracy, UA: User's Accuracy.

285

From the ten generated confusion matrices, the minimum Kappa value found with ML algorithm was 0.69, and the maximum value was 0.99, with a mean of 0.90 and a standard deviation of 0.094. For RF, the minimum Kappa value found was 0.85and the maximum was 0.99, with a mean of 0.94 and a standard deviation of 0.043.

A significant difference was verified by the Z test (Z > 1.96 at 95% of confidence level) (Table 5) among Kappa values for each classifier. In addition, RF algorithm presented the best Kappa and OA results, being considered the best alternative for image processing.

When validation was performed using the field samples, the Kappa indexes remained in the excellent range for both algorithms, 0.97 for ML and RF. The OA was also considered high (0.98 for both algorithms), proving that the field data are consistent with the results obtained through training and validation when using samples selected in the image itself.

The variables importance generated by the RF for UAV image, in decreasingorder, were red band, green, SAVI, NDVI and near infrared.

301

302 3.2. Classification using Maximum Likelihood and RF algorithms from LIDAR data
303 composition images

304 The images obtained by LIDAR data composition, classified using ML and RF305 algorithms is presented in Figure 4, and it represents part of the restoration project.

306

307



Fig. 4. a) LIDAR data composition images (Intensity, nDSM, DSM, DTM); b) Image
classified by Maximum Likelihood algorithm; c) Image classified by Random Forest
algorithm.

313 It is noticed that all the classes were well delimited on classified LIDAR images



According to Tables 3 and 4, Kappa and Overall Accuracy were high and ranked as excellent (Landis and Kock, 1977). User's and producer's accuracy values were also considered high for both algorithms indicating few confusions among classes. The greatest confusion was between "bare soil" and "grass cover" for both classifiers.

- 319
- 320
- 321

323 Table 3

324 Mean values generated from the ten confusion matrices, with the Maximum Likelihood 325 classifier, for LIDAR data images

CLASS	Bare soil	Grass cover	Canopy cover	Total	PA (%)
Bare soil	817	0	1	817	99.38
Grass cover	62	557	31	649	86.16
Canopy cover	72	8	678	758	89.50
Total	950	565	710	2225	
UA (%)	89.24	98.32	96.12		
Average Kappa				0.88	
Average OA				0.92	

326 OA: Overall Accuracy, PA: Producer's Accuracy, UA: User's Accuracy.

327

328 Table 4

329 Mean values generated from the ten confusion matrices, with the Random Forest 330 classifier, for LIDAR data images

CLASS	Bare soil	Grass cover	Canopy cover	Total	PA (%)
Bare soil	493	14	0	507	97.29
Grass cover	53	595	1	649	92.13
Canopy cover	0	0	642	642	100
Total	547	608	643	1798	
UA (%)	91.80	97.77	99.85		
Average Kappa				0.94	
Average OA				0.96	

331 OA: Overall Accuracy, PA: Producer's Accuracy, UA: User's Accuracy.

332

The minimum Kappa value found with ML was 0.61, and the maximum value was 0.98, with a mean of 0.88 and a standard deviation of 0.14. For RF, the minimum Kappa was 0.76, and the maximum was 0.98, with a mean of 0.94 and a standard deviation of 0.07.

337 A significant difference was verified by the Z test (Z > 1.96 at 95% of 338 confidence level) (Table 5) among the Kappa values for each classifier. RF algorithm 339 presented the best Kappa and OA results, then this algorithm was considered the best 340 alternative for LIDAR images processing.

When field samples were used to validate the classifications, the Kappa Indexremained in the excellent range, 0.86 for ML algorithm and 0.95 for RF. The OA was

also considered high, with 0.91 and 0.97 for ML and RF, respectively, proving that field
data are consistent with the results obtained through training and validation, using
samples collected in the image itself.

346 The variables importance generated by RF for LIDAR image, in decreasing347 order, were intensity band, nDMS, DSM and DTM.

348

349 3.3 Comparison between imaging methods: LIDAR and UAV camera

350 Considering that the best classification result was obtained using RF algorithm, 351 LIDAR and UAV methods were compared first for that algorithm. The difference 352 between LIDAR and UAV methods was nonsignificant (Z < 1.96 at 95% of confidence 353 level) (Table 5), demonstrating that method selection does not significantly affect the 354 results. On the other hand, when using ML, a significant difference (Z > 1.96 at 95% of 355 confidence level) (Table 5) was found between LIDAR and UAV images 356 classifications. In this case, the image from UAV camera produced a higher accuracy 357 classification.

358 Table 5 is a summary of Kappa and Z test values for each method evaluated on359 this study.

Table 5. Kappa Index found for UAV and LIDAR images classified by Maximum
Likelihood and Random Forest algorithms. The Table also presents the Z values for
comparison among the algorithm (RF and ML) for each imaging methods (UAV and
LIDAR), and the comparison between imaging methods (LIDAR and UAV) for each
algorithm (RF and ML)

Kappa Values	ML	RF	Z test ML x RF
UAV	0.90	0.94	14.55*
 LIDAR	0.88	0.94	5.74*
Z test UAV x LIDAR	2.58*	0.06	

³⁶⁵

366 * Significant difference (Z >1.96 at 95% of confidence level). LIDAR: Light Detection
367 and Ranging, ML: Maximum Likelihood; RF: Random Forest; UAV: Unmanned Aerial
368 Vehicle.

370 *3.4 Land cover composition obtained by classified images*

The area covered by canopy, grass and bare soil in the studied project generated through different image classification methods is represented in Figure 5. The highest percentage cover was generated for canopy cover followed by grass cover and bare soil in all classified images.



Fig. 5. Percentage area for canopy cover, grass cover and bare soil found through
different image classification methods: UAV camera images classified with ML and RF
algorithms, LIDAR images classified with ML and RF algorithms. LIDAR: Light
Detection and Ranging, ML: Maximum Likelihood; RF: Random Forest; UAV:
Unmanned Aerial Vehicle.

381

382 4. Discussion

We elaborated and compared methods of remote monitoring of forest restoration using LIDAR data and multispectral imaging of UAV camera and compared the efficiency of supervised classification algorithms ML and RF that proved to be reliable tools for assessing land cover composition. Field monitoring and evaluation of restoration requires a great deal of efforts mainly in large-scale projects (Zahawi et al., 2015; Reif and Theel, 2017; Ockendon et al, 2018), but the use of remote monitoring of

land cover indicators with synoptic, multispectral and multitemporal data (PanequeGalves et al., 2014; Reif and Theel, 2017) allow us to save time, fieldwork and
resources (Zahawi et al., 2015).

392

393 *4.1 Comparison between algorithms: Maximum Likelihood and Random Forest*

As previously mentioned, Kappa index and OA were chosen to evaluate the accuracy of the classification since those measures are still the most used in image classification, although they have recently been questioned (Lyons et al., 2018). Similar values of Kappa and OA were found in all classifications and these values were also high, denoting the classifier precision.

399 Although Kappa values were similar and ranked as excellent in all combinations 400 of images and algorithms, RF presented higher values of accuracy than the ML. This is 401 in accordance with studies carried out by Gislason et al. (2006), which RF efficiency 402 (Overall Accuracy = 82.8%) proved to be better when compared with other classifiers 403 (ensemble methods and CART - decision trees) for land cover classification with an 404 image composition (Landsat satellite images, elevation, slope and aspect) in Colorado. 405 RF algorithm has been widely used by ecologists because of its simple interpretation, 406 high accuracy, rapid processing, robustness to outliers and noises and ability to 407 characterize complex interactions between variables (Breiman, 2011; Cutler et al., 408 2007).

An interesting feature of RF classification is the possibility of obtaining the variables importance, especially in situations where it is necessary to classify hiperspectral images (Gislason et al., 2006). In our study with multispectral images, intensity band and nDSM were the most important variables of LIDAR image classification. This can be justified by the fact that intensity band has spectral

414 information and nDSM has height information, in other words, they are poorly
415 correlated, providing less redundant data for the classifier (Lu and Weng, 2007). For
416 UAV images, the most important variables were the red and green bands, followed by
417 SAVI index, which also are the less correlated variables of the composition.

418 Although the ML algorithm was slightly less assertive than the RF, it also 419 presented results ranked as excellent, which makes it eligible for this type of 420 monitoring. This classifier is widely used for remote sensing, displaying good results 421 when the data have a normal distribution and when sample selection represents well the 422 spectral diversity of the class to be mapped (de Oliveira et al., 2013). ML has already 423 proved efficient in several studies, such as the study by Silva et al. (2016), who tested 424 the efficiency of this classifier, after segmentation, for monitoring Brazilian Savanna 425 (*Cerrado*) land cover with UAV image, and obtained high values of similarity (0.94) 426 with the visual interpretation. In the case of de Oliveira et al. (2013), who mapped forest 427 fragments with monodominant aroeira (Myracrodruon urundeuva), ML classifier 428 presented the best performance (Kappa = 80) in RapidEye images classification when 429 compared to Neural Networks.

430

431 *4.2 Comparison between imaging methods: LIDAR and UAV camera*

No significant difference was found when comparing the sensors (LIDAR or
UAV camera) using RF algorithm. Therefore, regardless the method choice, good
results are expected to be obtained when RF is used. In case of ML, there was
significant difference between the methods, so UAV images presented better results
than LIDAR.

437 The "shadow" class was the one that mostly generated confusion in UAV438 camera images for both classifiers. This confusion can be justified by the fact that

439 shadows cause partial or total loss of radiometric signature in the analyzed areas thus 440 this can affect the performance of image classification process, object identification and 441 consequently the land-cover mapping (Adeline et al., 2013; Movia et al., 2016). Shadow 442 in images can be considered as a disadvantage of the use of UAV camera images when 443 compared to LIDAR. The latter is considered an active sensor that have their own light 444 source and does not require sunlight, so shadow interference does not occur in images 445 (Song et al, 2002; Giongo et al., 2010) reducing the probability of class confusions in 446 classification.

447 Little confusion was found between the classes when LIDAR data were 448 evaluated. In this case, the greatest confusion was between bare soil and grass cover. In 449 a study conducted by Song et al. (2002) using LIDAR, it was found that only with 450 intensity band it would be difficult to separate grass cover from canopy cover, but with 451 the addition of nDSM band, this becomes possible. According to Thenkabail et al. 2004, 452 an increase in spectral bands quantity may improve the accuracy of classification, but 453 only when these bands are useful for discriminating the classes. The low spectral 454 information of the analyzed images does not show an adverse influence on the 455 classifications since both Kappa and OA values were high, with few confusing errors. 456 However, selecting samples within LIDAR images was more difficult since the small 457 number of spectral bands hinder the visual differentiation among classes.

The use of LIDAR has been shown to be efficient for forest applications because it mobilizes many points with high precision, low cost and high speed of data acquisition (Giongo et al., 2010). Additionally, LIDAR data has already been proved to be successful (Overall Accuracy = 75%) for grassland habitats classification with RF algorithm as showed on the study developed by Zlinszky et al. (2014) in Hungary. According to data provided by Fibria and to Zahawi et al., 2015, the use of LIDAR is

464 more financially feasible for large areas. On the other hand, UAV is feasible for small465 area imaging (Paneque-Galves et al., 2014; Zahawi et al., 2015).

466 To achieve better results on remote monitoring and make the evaluation of a 467 greater number of indicators possible, such as arboreal individual density, species 468 richness, invasive tree cover and vegetation strata, it is suggested the use of aerial 469 cameras with high resolution and LIDAR sensor coupled, or Ecosynth (new aerial 470 remote sensing system with similar properties to LIDAR, but with RGB spectral 471 attributes for each point, e.g. Zahawi et al., 2015). In this case, in addition to increase 472 spectral information, there will be geometry data (Persson et al., 2004). Additionally, 473 the use of region-based image classification, through segmentation of images it is 474 suggested to improve accuracy values. In the study carried out by Holmgren et al. 475 (2008) in Scandinavia to identify local occurrence of species, higher value of OA (96%) 476 was found when segmentation was used before the classification of aerial camera 477 images combined with LIDAR data by the ML algorithm. The combination of both 478 devices allows us to evaluate a greater number of indications with high accuracy, but it 479 makes the monitoring process more expensive.

480

481 *4.3 Land cover composition by classified images*

Slight variations in the area occupied by different classes were observed between the image classification methods. These little variations may be associated with the presence of shaded areas in UAV camera images (cf. Adeline et al., 2013; Movia et al., 2016), that probably consist of bare soil or grass in the field or, additionally, by wrongly classified pixels in the digital image processing.

487 The highest cover was generated for canopy cover by all methods, followed by488 grass cover and then bare soil. According to the Atlantic Forest Restoration Pact (2013)

489 and Viani et al. (2017), after four years of restoration implementation the canopy cover 490 should achieve at least 70% to be capable to reduce invasive grass cover and to facilitate 491 tree establishment. Our study area has not reached the desired value yet, probably due to 492 the planting methodology, which used lines of early seral tree species interspersed with 493 late seral species that have slower development and higher mortality. The high grass 494 cover found in this study can be explained by the fact that the project area was covered 495 by degraded pastures dominated by Urochloa decumbens before restoration activities 496 and by the short time period since the restoration process has started. The invasive grass 497 cover also competes with native tree seedlings species impeding or hindering their 498 growth (Rocha-Nicoleite et al, 2017), and this can be seen in our restoration area. Bare 499 soil cover was not high, but when it occurs in large areas, it induces soil erosion and 500 nutrient loss (Muñoz-Rojas et al., 2016) that call for further restoration interventions.

501 According to the Atlantic Forest Restoration Pact (2013) and Viani et al. (2017), 502 after checking the land cover indicators, a field ecological monitoring (phase II) has to 503 be done in areas where canopy cover is greater than 70%. This monitoring is to verify 504 species composition indicators (e.g. richness, density, regeneration) based on field 505 sampling and it will guide adaptive management techniques aiming to correct the 506 restoration trajectory of the areas if necessary. This two-step evaluation (remote sensing 507 and field monitoring) should reduce the costs associated with monitoring, since it will 508 not be necessary to measure field indicators in all areas. The field study related to this 509 monitoring will be presented in a future paper.

510

511 5. Conclusions

512 The methods employed in this study are efficient to monitor restoration areas,513 bringing gains in quality and precision, synoptic analysis and reduction of field efforts,

514 especially on a large scale. RF algorithm presented the greatest assertiveness in the 515 image classification by UAV camera and LIDAR. The land cover composition found on 516 this research suggests that the study area has not achieved the restoration success yet, 517 therefore adaptive management strategies must be adopted to correct their trajectory 518 towards the desired target state for restoration projects.

519

520 Acknowledgments

521 The authors of this study thank Fibria Celulose S.A. for providing the data and
522 financial support and the National Council for Scientific and Technological
523 Development (CNPq) for granting the first author's Master's scholarship and research
524 productivity for the second author.

525

526 References

- Adeline, K. R. M., Chen, M., Briottet, X., Pang, S. K., Paparoditis, N., 2013. Shadow
 detection in very high spatial resolution aerial images: A comparative
 study. ISPRS Journal of Photogrammetry and Remote Sensing, 80, 21-38.
 DOI:10.1016/j.isprsjprs.2013.02.003
- Almeida, T.M., Moreau, A. M.S.S., Moreau, M.S., Pires, M.D.M., Fontes, E.D.O.,
 Góes, L. M., 2008. Reorganização socioeconômica no extremo sul da Bahia
 decorrente da introdução da cultura do eucalipto. Sociedade & Natureza, 20 (2),
 5-18. (Abstract in English). http://dx.doi.org/10.1590/S198245132008000200001
- 537 Atlantic Forest Restoration Pact (Pacto pela Restauração da Mata Atlântica). Protocolo
 538 de Monitoramento para Programas e Projetos de Restauração Florestal
 539 (Monitoring Protocol for Forest Restoration Projects). 2013. Available in:
 540 http://media.wix.com/ugd/5da841_c228aedb71ae4221bc95b909e0635257.pdf
 541 . Accessed in: 3 Jul. 2017.

- 542 Bortolot, Z.J., & Wynne, R.H., 2005. Estimating forest biomass using small footprint
 543 LIDAR data: An individual tree-based approach that incorporates training
 544 data. ISPRS Journal of Photogrammetry and Remote Sensing, 59 (6), 342-360.
 545 DOI:10.1016/j.isprsjprs.2005.07.001
- 546 Breiman, L. 2001. Random forests. Machine Learning, 45, 15–32.
- 547 Campbell, J. B., 1996. Introduction to remote sensing. New York: The Guilford Press,
 548 622 p.
- 549 Cutler, R., Edwards, T.C., Beard, K.H. Cutler, A., Hess, K.T., Gibson, J., Lawler, J.J.,
 550 2007. Random Forests for classification in ecology. Ecology, 88 (11), 2783551 2792. DOI:10.1890/07-0539.1
- 552 Costa, O.V., Cantarutti, R.B., Fontes, L.E.F., Costa, L.D., Nacif, P.G.S., Farias, J.C.,
 553 2009. Estoque de carbono do solo sob pastagem em área de tabuleiro costeiro no
 554 sul da Bahia. Revista Brasileira de Ciência do Solo, 33 (5), 1137-1145. (Abstract
 555 in English). http://dx.doi.org/10.1590/S0100-06832009000500007
- 556 Congalton, R., Green, K., 2009. Assessing the accuracy of remotely sensed data:
 557 Principles and Practices. Boca Raton: CRC Press, Taylor & Francis Group, 183
 558 p.
- 559 Dubayah, R.O., Drake, J.B., 2000. LIDAR Remote Sensing for Forestry. Journal of
 560 Forestry, 98, 44-46. https://doi.org/10.1093/jof/98.6.44
- 561 ESRI, Environmental Systems Research Institute, 2015. ArcGIS for Desktop. Version
 562 10.3. Redlands, California, United States.
- Foody, G.M., 2004. Thematic map comparison: evaluating the statistical significance of
 differences in classification accuracy. Photogrammetric Engineering & Remote
 Sensing, 70 (5), 627-633. http://dx.doi.org/10.14358/PERS.70.5.627
- 566 Getzin, S., Wiegand, K., Schoening, I., 2012. Assessing biodiversity in forests using
 567 very high-resolution images and Unmanned Aerial Vehicles. Methods in
 568 Ecology and Evolution. 3, 397-404. DOI: 10.1111/j.2041-210X.2011.00158.x
- Giongo, M., Koehler, H. S., do Amaral Machado, S., Kirchner, F.F., Marchetti, M.,
 2010. LIDAR: princípios e aplicações florestais. Pesquisa Florestal
 Brasileira, 30 (63), 231. (Abstract in English). DOI: 10.1111/j.2041210X.2011.00158.x
- 573 Gislason, P.O., Benediktsson, J. A., Sveinsson, J. R., 2006. Random forests for land
 574 cover classification. Pattern Recognition Letters, 27 (4), 294-300.
 575 DOI:10.1016/j.patrec.2005.08.011

- 576 Hagner, O., Reese, H., 2007. A method for calibrated maximum likelihood
 577 classification of forest types. Remote sensing of environment, 110(4), 438-444.
 578 https://doi.org/10.1016/j.rse.2006.08.017
- 579 Harris, C. J., Leishman, M. R., Fryirs, K. and Kyle, G., 2012. How Does Restoration of
 580 Native Canopy Affect Understory Vegetation Composition? Evidence from
 581 Riparian Communities of the Hunter Valley Australia. Restoration Ecology, 20:
 582 584–592. DOI:10.1111/j.1526-100X.2011.00823.x
- Hastie, T., Tibshirani, R., Friedman, J., 2009. The elements of statistical learning: data
 mining, inference, and prediction. Springer New York: Springer Series in
 Statistics, 758p.
- Holmgren J., Persson Å., Söderman U., 2008. Species identification of individual trees
 by combining high resolution LIDAR data with multi spectral images,
 International Journal of Remote Sensing, 29 (5), 1537-1552. DOI:
 10.1080/01431160701736471
- Huete, A.R. A soil adjusted vegetation index (SAVI), 1988. Remote Sensing of
 Enviroment, 25 (3), 295-309. https://doi.org/10.1016/0034-4257(88)90106-X
- 592 Ioki, K., Imanishi, J., Sasaki, T., Morimoto, Y., Kitada, K., 2010. Estimating stand
 593 volume in broad-leaved forest using discrete-return LIDAR: plot-based
 594 approach. Landscape and Ecological Engineering, 6 (1), 29-36. DOI:
 595 10.1007/s11355-009-0077-4
- 596 Ivanauskas, N. M., Assis, M.C., 2012. Formações florestais brasileiras. In: Martins,
- 597 S.V., (Org.). Ecologia de Florestas Tropicais do Brasil. 2ed.Viçosa-MG: Editora
- **598** UFV, v. 1, p. 01-371.
- Kim, S., McGaughey, R.J., Andersen, H.E., Schreuder, G., 2009. Tree species
 differentiation using intensity data derived from leaf-on and leaf-off airborne
 laser scanner data. Remote Sensing of Environment, 113 (8), 1575-1586. DOI:
 10.1016/j.rse.2009.03.17
- Koh, L. P., Wich, S. A., 2012. Dawn of drone ecology: low-cost autonomous aerial
 vehicles for conservation. Tropical Conservation Science, 5(2):121-132.
 Available online: www.tropicalconservationscience.org
- 606 Kopen
- Landis, J., Koch, G.G., 1977. The measurements of agreement for categorical data.
 Biometrics, 33 (1), 159-174. DOI: 10.2307/2529310

- Lawrence, R.L., Wood, S.D., Sheley, R.L., 2006. Mapping invasive plants using hyperspectral imagery and Breiman Cutler classifications (Random Forest). Remote
 Sensing of Environment, 100, 356–362. DOI: 10.1016/j.rse.2005.10.014
- 612 Lefsky, M. A., Cohen, W. B., Parker, G.G., Harding, D. J., 2002. LIDAR remote
 613 sensing for ecosystem studies. BioScience, v. 52, n. 1, p. 19-30.
 614 http://dx.doi.org/10.1641/0006-3568(2002)052[0019: LRSFES]2.0.CO;2
- Li, T., Lü, Y., Fu, B., Comber, A. J., Harris, P., & Wu, L. (2017). Gauging policydriven large-scale vegetation restoration programmes under a changing
 environment: Their effectiveness and socio-economic relationships. Science of
 the Total Environment, 607, 911-919.
 http://dx.doi.org/10.1016/j.scitotenv.2017.07.044
- Lu, D., Weng Q., 2007. A survey of image classification methods and techniques for
 improving classification performance. International Journal of Remote Sensing,
 28 (5), 823-870. DOI: 10.1080/01431160600746456
- Lyons, M. B., Keith, D. A., Phinn, S. R., Mason, T. J., Elith, J., 2018. A comparison of
 resampling methods for remote sensing classification and accuracy
 assessment. Remote Sensing of Environment, 208, 145-153.
- Mascaro, J., Asner, G.P., Davies, S.J., Dehgan, A., Saatchi, S., 2014. These are the days
 of lasers in the jungle. Carbon Balance Management, 9, (1),
 7. DOI: 10.1186/s13021-014-0007-0
- Melo, F.P.L., Pinto, S.R.R., Brancalion, P.H.S., Castro, P.S., Rodrigues, R.R., Aronson,
 J., Tabarelli, M., 2013. Priority setting for scaling-up tropical forest restoration
 projects: early lessons from the Atlantic Forest Restoration Pact. Environmental
 Science and Policy, 33, 395–404. http://dx.doi.org/10.1016/j.envsci.2013.07.013
- Michez, A., Piégay, H., Jonathan, L., Claessens, H., & Lejeune, P., 2016. Mapping of
 riparian invasive species with supervised classification of Unmanned Aerial
 System (UAS) imagery. International Journal of Applied Earth Observation and
 Geoinformation, 44, 88-94. http://dx.doi.org/10.1016/j.jag.2015.06.014
- Moreau, A.M.S.S., Ker, J.C., Costa, L.M., Gomes, F.H., 2006. Caracterização de solos
 de duas toposequências em Tabuleiros Costeiros do Sul da Bahia. Revista
 Brasileira de Ciência do Solo, 30, 1007-1019. (Abstract in English)
 http://dx.doi.org/10.1590/S0100-06832006000600010
- 641 Movia, A., Beinat, A., Crosilla, F., 2016. Shadow detection and removal in RGB VHR
 642 images for land use unsupervised classification. ISPRS Journal of

- 643
 Photogrammetry
 and
 Remote
 Sensing, 119,
 485-495.

 644
 https://doi.org/10.1016/j.isprsjprs.2016.05.004

 <td
- Muñoz-Rojas, M., Erickson, T. E., Dixon, K. W. and Merritt, D. J., 2016. Soil quality
 indicators to assess functionality of restored soils in degraded semiarid
 ecosystems. Restor Ecol, 24: S43–S52. doi:10.1111/rec.12368
- de Oliveira, F.P., Fernandes Filho, E. I., Soares, V.P., 2013. Mapeamento de fragmentos
 florestais com monodominância de aroeira a partir da classificação
 supervisionada de imagens RapidEye1. Revista Árvore, 37 (1), 151-161.
 (Abstract in English) http://dx.doi.org/10.1590/S0100-67622013000100016.
- 652 Ockendon, N., Thomas, D.H.L., Cortina, J., Adams, W.M., Aykroyd, T., Barov, B., 653 Boitani, L., Bonn, A., Branquinho, C., Brombacher, M., Burrellm, C., Carvern, 654 S., Cricko, H.Q.P., Duguyp, B., Everettq, S., Fokkensr, B., Fullers. 655 R.J., Gibbonst, D. W., Gokhelashviliu, R., Griffinv, C., Halleyw, J.D., Hothamx, 656 P., Hughesy, F.M.R., Karamanlidisz, A.A., McOwenaa, C.J., Milesaa, 657 L., Mitchellab, R., Randsac, M.R.W. Robertsad, J., Sandomae, C.J., Spenceraf, 658 J.W, Broekeag. E., Tewb, E.R., Thomasah, C.D., Anastasiya Timoshynaai, A., 659 Unsworthaj, R.K.F., Warringtonak, S., Sutherlandb, W.J., 2018. One hundred 660 priority questions for landscape restoration in Europe. Biological Conservation, 661 221, 198-208. https://doi.org/10.1016/j.biocon.2018.03.002
- Paneque-Gálvez, J., McCall, M.K., Napoletano, B.M., Wich, S.A., & Koh, L.P, 2014.
 Small drones for community-based forest monitoring: An assessment of their
 feasibility and potential in tropical areas. Forests, 5 (6), 1481-1507. DOI:
 10.3390/f5061481
- 666 Persson, A., Holmgren, J., Söderman, U., Olsson, H., 2004. Tree species classification 667 of individual trees in Sweden by combining high resolution laser data with high 668 resolution near-infrared digital images. International Archives of 669 Photogrammetry, Remote Sensing and Spatial Information Sciences, 36 (8), 670 204-207.
- Puissant, A., Rougier, S., Stumpf, A., 2014. Object-oriented mapping of urban trees
 using Random Forest classifiers. International Journal of Applied Earth
 Observation and Geoinformation, 26, 235-245.
 http://dx.doi.org/10.1016/j.jag.2013.07.002

- 675 R Core Team, 2016. R: A language and environment for statistical computing. R
 676 Foundation for Statistical Computing, Vienna, Austria. URL http://www.R677 project.org/.
- Reif, M.K., Theel, H.J., 2017. Remote sensing for restoration ecology: application for
 restoring degraded, damaged, transformed, or destroyed ecosystems Integr.
 Environ. Assess. Manag., 13, 614–630. DOI 10.1002/ieam.1847
- Rouse, J.W., Haas, R.H., Schell, J.A., Deering, D.W., 1973. Monitoring vegetation
 systems in the Great Plains with ERTS. In3rd ERTS Symposium, NASA SP-351
 I, 309–317.
- Rocha-Nicoleite, E., Overbeck, G.E., Müller, S.C., 2017. Degradation by coal mining
 should be priority in restoration planning. Perspect. Ecol. Conserv. 15, 197–200.
 https://doi.org/10.1016/j.pecon.2017.05.006
- Rodrigues, R.R., Lima, R.A., Gandolfi, S., Nave, A. G., 2009. On the restoration of
 high diversity forests: 30 years of experience in the Brazilian Atlantic Forest.
 Biological conservation, 142(6), 1242-1251. DOI:10.1016/j.biocon.2008.12.008
- Ruiz-Jaen, M.C., Aide, T.M., 2005. Restoration Success: How is it being measured?
 Restoration Ecology, 13, 569 -577. DOI: 10.1111/j.1526-100X.2005.00072.x
- Saporetti Junior, A.W., Schaefer, C.E.G.R., Souza, A.L., Soares, M.P., Araújo, D.S.D.,
 Meira-Neto J.A.A., 2012. Influence of soil physical properties on plants of
 Mussununga ecosystem, Brazil. Folia Geobot, 47, 29-39. DOI 10.1007/s12224011-9106-9
- 696 Silva, F.C.M., Silva, N. M., Cândido, A.K.A.A., 2016. Seleção de técnicas de
 697 classificação de fotografias aéreas derivadas de VANT na análise ambiental de
 698 área de cerrado. REDE-Revista Eletrônica do PRODEMA, 10 (1), 74-84.
 699 (Abstract in English)
- Song, J. H., Han, S. H., Yu, K. Y., Kim, Y.I., 2002. Assessing the possibility of landcover classification using LIDAR intensity data. International Archives of
 Photogrammetry Remote Sensing and Spatial Information Sciences, 34 (3/B),
 259-263.
- Thenkabail, P.S., Enclona, E.A., Ashton, M.S., Legg, C. and De Dieu, M.J., 2004.
 Hyperion, IKONOS, ALI, and ETM+ sensors in the study of African rainforests.
 Remote Sensing of Environment, 90, 23-43. DOI:10.1016/j.rse.2003.11.018

- 707 Veloso, H.P., Rangel Filho, A.L.R., Lima, J.C.A., 1991. Classificação da vegetação
 708 brasileira adaptada a um sistema universal. Rio de Janeiro: IBGE, Departamento
 709 de Recursos Naturais e Estudos Ambientais. 124 p.
- Viani, R.A., Holl, K.D., Padovezi, A., Strassburg, B.B., Farah, F.T., Garcia, L.C.,
 Chaves, R.B., Rodrigues, R.R., Brancalion, P. H., 2017. Protocol for monitoring
 tropical forest restoration: perspectives from the Atlantic forest restoration pact
 in Brazil. Tropical Conservation Science, 10,
 https://doi.org/10.1177/1940082917697265
- Xian, G., Homer, C., Fry, J., 2009. Updating the 2001 National Land Cover Database
 land cover classification to 2006 by using Landsat imagery change detection
 methods. Remote Sensing of Environment, 113(6), 1133-1147.
 DOI:10.1016/j.rse.2009.02.004
- Zahawi, R.A., Dandois, J.P., Holl, K.D., Nadwodny, D., Reid, J. L., Ellis, E. C., 2015.
 Using lightweight unmanned aerial vehicles to monitor tropical forest
 recovery. Biological Conservation, 186, 287-295.
 http://dx.doi.org/10.1016/j.biocon.2015.03.031
- Zlinszky, A., Schroiff, A., Kania, A., Deak, B., Mücke, W., Vari, A., Szekely,
 B., Pfeifer, N., 2014. Categorizing grassland vegetation with full-waveform
 airborne laser scanning: a feasibility study for detecting Natura 2000 habitat
 types. Remote Sens., 6, 8056-8087. DOI:10.3390/rs6098056
- Zonete, M.F., Rodriguez, L.C.E., Packalén, P., 2010. Estimação de parâmetros
 biométricos de plantios clonais de eucalipto no sul da Bahia: uma aplicação da
 tecnologia laser aerotransportada. Scientia Forestalis, 38 (86), 225-235.
 (Abstract in English)