



Title	Prenatal organochlorine pesticide exposure and the disruption of steroids and reproductive hormones in cord blood : The Hokkaido study
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Citation	Environment international, 110, 1-13 https://doi.org/10.1016/j.envint.2017.10.006
Issue Date	2018-01
Doc URL	http://hdl.handle.net/2115/76437
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Type	article (author version)
File Information	ENVINT_2017_824_(final).pdf



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1 **Prenatal organochlorine pesticide exposure and the disruption of steroids and**
2 **reproductive hormones in cord blood: The Hokkaido Study**

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24 **Word count:** 4882

60
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64 25 ***Number tables:*** 6
65
66 26 ***Number of figures:*** 1
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91 38 Conflicts of interest: none
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123 41 **Abstract**

124 42 Certain organochlorine pesticides (OCPs) are designated as persistent organic pollutants
125 43 and are regulated in many countries. The effects of OCPs on pediatric endocrinology are
126 44 a concern; however, only limited data exist from human studies on maternal OCP
127 45 exposure and its effects on infants' hormone levels. This study was conducted as part of
128 46 the Hokkaido Study Sapporo Cohort, a prospective birth cohort study in Japan.
129
130 47 Participants included 514 women who enrolled at 23–35 weeks of gestation between
131 48 2002 and 2005; maternal blood samples were collected in late pregnancy, and 29 OCPs
132 49 were measured. Reproductive and steroid hormone levels in cord blood were also
133 50 determined. Characteristics of mothers and their infants were obtained from self-
134 51 administered questionnaires and medical records. Ultimately, 232 samples with both
135 52 OCP and hormone data were analyzed. Fifteen of 29 investigated OCPs were detected
136 53 in over 80% of the samples, with *p,p'*-dichlorodiphenyldichloroethylene showing the
137 54 highest concentration (median value: 619 pg/g-wet). The association between OCPs and
138 55 sex hormone levels varied by sex. Linear regression models after sex stratification
139 56 showed that chlordanes, cis-hexachlorobenzene, heptachlor epoxide, Mirex, and
140 57 toxaphenes in maternal blood were inversely associated with testosterone, cortisol,
141 58 cortisone, sex hormone-binding globin, prolactin, and androstenedione-
142 59 dehydroepiandrosterone (DHEA) and testosterone-androstenediones ratios among boys.
143 60 Furthermore, these OCPs were positively correlated with DHEA, follicle stimulating
144 61 hormone (FSH), and adrenal androgen-glucocorticoid and FSH-inhibin B ratios among
145 62 boys. In categorical quartile models, testosterone and DHEA were inversely and
146 63 positively associated with OCPs, respectively. Estradiol-testosterone and adrenal
147 64 androgen-glucocorticoid ratios tended to increase with increasing OCP concentrations
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182 65 in the higher quartile, while the testosterone-androstenedione ratio tended to decrease.
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184 66 Sex hormone-binding globulin and prolactin showed an inverse association with OCPs.
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186 67 Among girls, the linear regression model showed that only *p,p'*-
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188 68 dichlorodiphenyltrichloroethane was inversely associated with the level of DHEA and
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190 69 the adrenal androgen-glucocorticoid ratio, but was positively associated with cortisone
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192 70 levels. However, no associations were observed using the quartile categorical model.
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194 71 These results suggest that prenatal exposure to OCPs disrupt reproductive hormones of
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196 72 fetuses in utero among boys, even at relatively low levels.

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202 74 **Key Words:** Organochlorine pesticides; reproductive hormones; steroid hormones;
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204 75 prenatal exposure; cord blood; birth cohort

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208 77 **Abbreviations:**
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210 78 CI, confidence interval
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212 79 CYP11A1, cytochrome P450 family 11 subfamily A member 1
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214 80 CYP17A1, cytochrome P450 family 17 subfamily A member 1
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216 81 CYP19A1, cytochrome P450 family 19 subfamily A member 1
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218 82 DDD, dichlorodiphenyldichloroethane
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220 83 DDE, dichlorodiphenyldichloroethylene
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222 84 DDT, dichlorodiphenyltrichloroethane
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224 85 HCB, hexachlorobenzene
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226 86 HCE, heptachlor epoxide
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228 87 HCH, hexachlorocyclohexane
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230 88 HSD17B1, hydroxysteroid 17-beta dehydrogenase 1

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241 89 HSD3B1, hydroxy-delta-5-steroid dehydrogenase, 3 beta- and steroid delta-isomerase 1
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243 90 IRMA, immunoradiometric assay
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245 91 DHEA, dehydroepiandrosterone
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247 92 EIA, enzyme immunoassay
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249 93 ELISA, enzyme-linked immunosorbent assay
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251 94 FSH, follicle stimulation hormone
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253 95 INSL3, insulin-like factor 3
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255 96 LC-MSMS, liquid chromatography-tandem mass spectrometry
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257 97 LH, luteinizing hormone
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259 98 LSM, least square mean
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261 99 OCP, organochlorine pesticides
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263 100 SHBG, sex hormone-binding globulin
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265 101 StAR, steroidogenic acute regulatory protein
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300 103 **1. Introduction**
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303 104 Organochlorine pesticides (OCPs) are chlorinated hydrocarbons used extensively in the
304 105 1940s for agriculture and pesticide control, and are now designated as persistent organic
306 106 pollutants by the Stockholm Convention (<http://chm.pops.int>). Although the Stockholm
307 107 Convention has issued an exemption for the production and public health use of
308 108 dichlorodiphenyltrichloroethane (DDT) to control vector-borne diseases, most OCPs
309 109 were banned in the United States, Europe, and many other countries in the early 1970s
310 110 (WHO, 2012). The use of OCPs has been eliminated or restricted in Japan since the
311 111 1970's (Kanazawa et al. 2012). Although most OCPs have been prohibited for over 30
312 112 years, they are still detected in the environment and in human populations. According to
313 113 Japanese monitoring data, the levels of DDT and its metabolites in water and sediment
314 114 have decreased since 1990 and have consistently remained low since 2000; however,
315 115 they are still detectable (Ministry of Environment, Japan 2006). Heptachlor epoxide
316 116 (HCE), hexachlorocyclohexane (HCH), Mirex, Parlar-26, and Parlar-50 are also above
317 117 detectable levels in water and sediments, even though the latter three have never been
318 118 used in Japan (Ministry of Environment, Japan 2006).

319 119 The endocrine disrupting properties of OCPs are considered a health concern. In
320 120 previous cross-sectional studies among adults, heptachlor and *o,p'*-DDT concentrations
321 121 were associated with lower testosterone levels in men (Freire et al. 2014). In women,
322 122 hexachlorobenzene (HCB), *p,p'*-DDT, *p,p'*-dichlorodiphenyldichloroethane (DDD),
323 123 endosulfan, aldrin, and Mirex showed inverse associations with luteinizing hormone
324 124 (LH) and follicle stimulation hormone (FSH) while showing positive associations with
325 125 prolactin (Freire et al. 2014).

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359 126 Maternal exposure to OCPs may affect fetal hormone levels. Sex steroid
360 127 hormones including testosterone, progesterone, and estradiol exert their functions
361 128 predominantly in the gonads, and dehydroepiandrosterone (DHEA) and
362 129 androstenedione are activated to form androgens and estrogens that have important roles
363 130 in sex differentiation and maturation (Labrie et al. 2001). Cortisol and cortisone are
364 131 synthesized within the adrenal cortex, are involved in a wide range of physiological
365 132 processes, and are essential for regulating and/or modulating homeostasis in
366 133 metabolism, growth, neurodevelopment, and the immune system (Braun et al. 2013;
367 134 Reynolds 2010). LH and FSH play critical roles in the development and regulation of
368 135 numerous body functions via the hypothalamic-pituitary-gonadal (HPG) axis (Kuiri-
369 136 Hänninen et al. 2014). Inhibin B and insulin-like factor-3 (INSL3) are major products
370 137 secreted by the Leydig and Sertoli cells, respectively, and the establishment of sufficient
371 138 numbers of these cells is critical for the production of sperms in adulthood (Ivell et al.
372 139 2013; Orth and Boehm 1990). In response to gonadotropins, testosterone (via LH
373 140 signaling) and inhibin B together act to regulate the secretion of FSH; these constitute
374 141 the major negative feedback signals that maintain the physiological function of the HPG
375 142 axis (Carlson, 2009). However, only limited data exist regarding human studies on
376 143 prenatal exposure to OCPs and their effects on steroids and reproductive hormone levels
377 144 in offspring. There is only one study in France that found that prenatal α -endosulfan and
378 145 HCE increase estradiol and sex hormone-binding globulin (SHBG), whereas these same
379 146 agents reduce testosterone levels at birth (Warembourg et al. 2016).
380
381 147 We have previously reported that 21 of 29 tested OCPs were detected in
382 148 maternal blood acquired between 2002 and 2005 in Japan (Kanazawa et al. 2012). The
383 149 impact of relatively low levels of OCP exposure on hormones at birth has still not been
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418 150 well-investigated in epidemiological studies. In particular, the effects of OCPs other
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420 151 than DDTs are rarely investigated. Thus, we hypothesized that prenatal exposure to
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422 152 even relatively low levels of these agents may alter hormone levels in infants. To that
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424 153 end, the aim of this study was to examine the associations between prenatal OCP
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426 154 exposure and cord blood steroid and reproductive hormone levels.
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431 156 **2. Methods**
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433 157 *2.1 Participants*
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435 158 This investigation was based on the Sapporo Cohort of the Hokkaido Study on
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437 159 Environment and Children's Health. Details of this study, including the population, data
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439 160 collection, sampling of the biological specimens, and contents of the administered
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441 161 questionnaire, were described previously (Kishi et al. 2017; Kishi et al. 2013; Kishi et
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443 162 al. 2011). Briefly, Japanese pregnant women who lived in Sapporo City or surrounding
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445 163 areas were recruited at 23–35 weeks of gestation between July 2002 and October 2005
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447 164 at an obstetrics and gynecology hospital in Sapporo, Hokkaido, Japan. Among the 1796
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449 165 eligible women approached, 25% were excluded because they were enrolled in the
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451 166 Japanese Cord Blood Bank or planned to deliver at another hospital. Ultimately, 514
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453 167 pregnant women (28.6% of those approached) were enrolled in this study.
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459 169 *2.2 OCP measurement*
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461 170 Maternal blood samples were obtained at the time of patients' hospital examinations
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463 171 following recruitment (n=296). If a blood sample could not be obtained during
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465 172 pregnancy because of maternal anemia, a sample was collected during post-partum
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467 173 hospitalization within a week after delivery (n=130). All samples were stored at –80°C
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478 174 until analysis. OCPs in whole blood were measured by gas chromatography/high-
479 resolution mass spectrometry and gas chromatography/negative-ion chemical-ionization
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481 175 mass spectrometry at IDEA Consultants, Inc. (Shizuoka, Japan) The 29 OCPs evaluated
482
483 176 in this study were 5 chlordanes (*cis*-chlordan, *trans*-chlordan, *cis*-nonachlor, *trans*-
484
485 177 nonachlor, and oxychlordan), 6 DDTs (*o,p'*-DDT, *p,p'*-DDT, *o,p'*-DDE, *p,p'*-DDE,
486
487 178 *o,p'*-DDD, and *p,p'*-DDD), 3 ‘drins’ (aldrin, dieldrin, and endrin), 3 heptachlors
488
489 179 (heptachlors, *cis*- HCE, and *trans*-HCE), HCB, 4 HCH isomers (α -HCH, β -HCH, γ -
490
491 180 HCH, and δ -HCH), Mirex, and 6 toxaphenes (Parlar-26, Parlar-41, Parlar-40, Parlar-44,
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493 181 Parlar-50, and Parlar-62). Details of the measurement methods have been described
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495 182 previously (Kanazawa et al. 2012).

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501 185 *2.3 Measurement of steroids and reproductive hormones*
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503 186 The methods used to measure steroids and reproductive hormones were described
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505 187 previously (Araki et al. 2017; Araki et al. 2014; Goudarzi et al. 2016). Briefly, the
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507 188 concentrations of 7 steroid hormones including progesterone, estradiol, testosterone,
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509 189 DHEA, androstanedione, cortisol, and cortisone in cord blood were measured using
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511 190 liquid chromatography-tandem mass spectrometry (LC-MSMS) (Yamashita et al.
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513 191 2007a; Yamashita et al. 2007b). An immunoradiometric assay (IRMA) was used to
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515 192 measure the concentrations of LH, FSH, and prolactin (Spac-S LH Kit, Spac-S FSH Kit,
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517 193 and Spac-Prolactin Kit, respectively, TFB, Inc., Tokyo Japan). SHBG was also
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519 194 measured using IRMA-Count SHBG (Siemens, Berlin, Germany). Concentrations of
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521 195 inhibin B were measured by using an enzyme-linked immunosorbent assay (ELISA)
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523 196 (Inhibin B Gen ELISA, Beckman Coulter, Inc., CA, USA), while INSL3 was measured
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525 197 by using an enzyme immunoassay (EIA) (INSL3/RLF [human] EIA kit, Phoenix

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536 198 Pharmaceutical., Inc., CA, USA). All hormone measurements were conducted at Aska
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538 199 Pharma Medical Co., Ltd (Kanagawa, Japan).

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542 201 *2.4 Questionnaires and medical records*

545 202 The participants completed a self-administered questionnaire that extracted information
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547 203 on maternal age, education level, household income, maternal smoking and alcohol
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549 204 consumption during the first trimester, and medical history. Information at the time of
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551 205 delivery, including pre-pregnancy body mass index, pregnancy complications,
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553 206 gestational age, infant sex, parity, congenital anomalies such as hypospadias and
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555 207 cryptorchidism, and infant size was obtained from medical records (Kishi et al. 2013;
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557 208 Kishi et al. 2011).

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561 210 *2.5 Statistical analyses*

564 211 Of the 514 participants, 10 were excluded from the study owing to miscarriage,
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566 212 stillbirth, relocation, or voluntary withdrawal prior to delivery. Among 426 maternal
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568 213 blood samples, 379 were of sufficient quantity for OCP analysis, while hormone
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570 214 measurements were obtained from 295 infant cord blood samples. Ultimately, 232
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572 215 matched maternal serum and cord blood samples (for OCP and hormone levels
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574 216 measurements, respectively) were included in the statistical analysis.

577 217 Associations between maternal OCP concentrations and infant steroid hormone
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579 218 levels were examined for each OCP separately via linear regression analysis. In each
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581 219 model, the OCP was the independent variable while the hormone was the dependent
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583 220 variable. Initially, linear regression models for both sexes combined were constructed
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585 221 with the interaction terms of sex × OCP levels added in each model; this revealed

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596 222 significant differences between the sexes. Next, the models were applied following
597 223 stratification by sex. OCP levels and the concentrations of steroid hormones were
598 224 converted to a \log_{10} scale because they were not normally distributed. Two-sided P -
599 225 values <0.05 were considered statistically significant. Selected OCPs with P -values
600 226 <0.1 in linear regression models were then categorized into concentration quartiles to
601 227 examine dose-response relationships. The interquartile range for each OCP
602 228 concentration and the least squares means (LSM) of log-transformed hormone levels
603 229 were calculated and back-transformed. To calculate a P -value for the trend, linear
604 230 contrast coefficients of $-3, -1, +1$, and $+3$ were assigned to the first, second, third, and
605 231 fourth quartiles, respectively (Goudarzi et al., 2016; Itoh et al., 2016). P -values for trend
606 232 <0.05 were considered statistically significant. The OCP levels in the first quartile were
607 233 also compared to those in the second, third, and fourth quartiles using the Dunnett-Hsu
608 234 method; P -values were adjusted using Bonferroni's correction ($P<0.0167$). When below
609 235 their detection limits, the half values of these detection limits were used for both OCPs
610 236 and hormones.

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612 237 In addition to single hormone levels, we also examined the product-to-substrate
613 238 ratios of hormones that are adjacent in the metabolic pathway to determine their enzyme
614 239 activity indices (Hicks et al. 2014). For example, the estradiol-testosterone ratio
615 240 represents the index of the cytochrome P450 family 19 subfamily A member 1
616 241 (CYP19A1), better known as aromatase. The androstenedione-DHEA ratio represents
617 242 the hydroxy-delta-5-steroid dehydrogenase, 3 beta- and steroid delta-isomerase 1
618 243 (HSD3B1) index; the testosterone-androstenedione ratio represents the hydroxysteroid
619 244 17-beta dehydrogenase 1 (HSD17B1) index; and the cortisone-cortisol ratio represents
620 245 the HSD3B1 index. Increasing and decreasing ratios suggest the up- and

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246 downregulation of enzyme activity, respectively. Additionally, the adrenal androgen
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656 (the sum of DHEA and androstenedione)-glucocorticoid (sum of cortisol and cortisone)
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658 ratio was examined to determine the balance shift of adrenal androgen (C19-steroids)
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660 and glucocorticoid (C21-steroids) (Goudarzi et al. 2016). Similarly, testosterone-LH
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662 and FSH-inhibin B ratios were examined as indices of gonadal function. The inclusion
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664 of covariates was examined based on biological considerations, and included maternal
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666 age (continuous), parity (primipara or multipara), and gestational age (continuous). All
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668 statistical analyses were performed using the Japanese version of IBM SPSS Statistics
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670 19 (IBM Analytics, NC, USA) and the Japanese version of JMP Pro 12 (SAS Institute
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672 Inc., NC, USA).

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676 257 *2.6 Ethical approval*

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678 This study was approved by the Institutional Ethical Board for Epidemiological Studies
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680 at Hokkaido University Graduate School of Medicine and Hokkaido University Center
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682 for Environmental and Health Sciences, in accordance with the principles of the
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684 Declaration of Helsinki. All participants provided written informed consent.

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688 263 **3. Results**

689
690 Table 1 shows the characteristics of the participants included in this study as well as
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692 those of the original cohort. Compared to the original cohort, the mean birth weight and
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694 gestational age were slightly larger among the participants; the vaginal delivery rate
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696 among participants was 99.1%. One infant with cryptorchidism was included in the
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698 study.

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713 269 The concentrations of OCP in maternal blood samples are shown in Table 2.
714
715 270 There were 15 OCPs (3 chlordanes [*cis*-nonachlor, *trans*-nonachlor, and oxychlordane],
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717 271 5 DDTs [*o,p'*-DDT, *p,p'*-DDT, *o,p'*-DDE, *p,p'*-DDE, *p,p'*-DDD], dieldrin, *cis*-HCE,
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719 272 HCB, β -HCH, Mirex, and 2 toxaphenes [Parlar-26 and Parlar-50]) that were above 80%
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721 273 of their detection limits; these OCPs were subjected to further analysis. The median
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723 274 concentration of *p,p'*-DDE was the highest at 619.26 pg/g-wet, followed by β -HCH at
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725 275 154.31 pg/g-wet, and HCH at 103.99 pg/g-wet. The distributions of OCPs in the
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727 276 original cohort are shown in Supplemental Table S1; *trans*-chlordanne, *cis*-nonachlor,
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729 277 and *trans*-nonachlor levels of participants in this study were slightly higher compared
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731 278 with the levels in those who were excluded (Mann-Whitney U test, $P=0.06$, 0.039, and
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733 279 0.040, respectively; data not shown). Associations between OCP levels and maternal
734
735 280 and child characteristics are shown in Supplemental Table S1. The levels of most OCPs
736
737 281 increased with the ages of the mothers. OCP concentrations were not significantly
738
739 282 associated with infant characteristics such as sex, birth weight, or gestational age.

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741 283 The steroid and reproductive hormone levels in infants are shown in Table 3.
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743 284 Testosterone and DHEA levels were significantly higher and lower in boys than in girls,
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745 respectively. The detection rate of LH, FSH, and inhibin B was below 30% in girls,
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747 285 among whom only 16 samples were analyzed for INSL3. Therefore, no further tests of
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749 286 LH, FSH, inhibin B, and INSL3 were conducted for girls. The distribution of all
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751 287 measured hormone levels are shown in Supplemental Table S3; there were no
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753 288 differences between subjects included and excluded from the study.

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755 289 The associations between OCPs and steroid and reproductive hormone levels for
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757 290 both sexes combined were examined (Supplemental Table S4). Sex-specific OCP
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759 291 interactions ($P<0.05$) between one or more hormones and *cis*-nonachlor, *trans*-

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772 293 nonachlor, *p,p'*-DDE, *o,p'*-DDT, dieldrin, and Mirex were observed, suggesting that the
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774 294 effects of OCPs differ according to sex.

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776 295 Adjusted regression coefficients (β) and 95% confidence intervals (CIs) for the
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778 296 association between a 10-fold increase of OCP and \log_{10} -transformed hormone levels as
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780 297 determined by a linear regression model in boys and girls are shown in Tables 4 and 5,
781
782 298 respectively. Among infant boys, Mirex was inversely associated with testosterone, and
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784 299 *p,p'*-DDE showed a positive association with the estradiol-testosterone ratio.

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786 300 Oxychlordane, *cis*-nonachlor, *trans*-nonachlor, dieldrin, *cis*-HCE, HCB, Mirex, Parlar-
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788 301 26, and Parlar-50 were positively associated with DHEA. Oxychlordane, *trans*-
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790 302 nonachlor, *cis*-HCE, and Mirex were inversely associated with the androstenedione-
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792 303 DHEA ratio. Oxychlordane, *cis*-nonachlor, *trans*-nonachlor, dieldrin, *cis*-HCE, HCB,
793
794 304 Mirex, and Parlar-50 showed inverse associations with the testosterone-androstenedione
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796 305 ratio. Both *trans*-nonachlor and Mirex were inversely associated with cortisol and
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798 306 cortisone. Oxychlordane, *cis*-nonachlor, *trans*-nonachlor, *cis*-HCE, Mirex, and Parlar-
800
801 307 50 were positively associated with the adrenal androgen-glucocorticoid ratio. β -HCH
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803 308 was inversely associated with SHBG, and β -HCH, Mirex, Parlar-26, and Parlar-50 were
804
805 309 positively associated with FSH. *o,p'*-DDE, *p,p'*-DDE, *o,p'*-DDT, *p,p'*-DDT, dieldrin, β -
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807 310 HCH, Mirex, and Parlar-50 were all inversely associated with prolactin. Finally, *cis*-
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809 311 nonachlor, *p,p'*-DDE, *p,p'*-DDT, *cis*-HCE, HCB, β -HCH, Mirex, Parlar-26, and Parlar-
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811 312 50 were inversely associated with the FSH-inhibin B ratio. There was no statistically
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813 313 significant association between OCPs and progesterone, estradiol, androstenedione,
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815 314 FSH, inhibin B, or INSL3. Among girls, *p,p'*-DDD was inversely associated with
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817 315 DHEA and the adrenal androgen-glucocorticoid ratio, and was positively associated

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831 316 with cortisone. There was no association between OCPs and progesterone, estradiol,
832
833 317 testosterone, androstenedione, cortisol, SHBG, or prolactin among girls.
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835
836 318 Figure 1 shows the relationships between hormones and OCPs in quartile models
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838 319 for boys (only in cases where $P < 0.05$ was observed). Testosterone showed a decreasing
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840 320 trend in relation to quartiles of Mirex, while LSM analysis of the estradiol-testosterone
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842 321 ratio showed an increasing trend of p,p' -DDE in relation to quartiles. DHEA showed
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844 322 increasing trends in relation to the quartiles of *cis*-Nonachlor, Dieldrin, Parlar-26, and
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846 323 Parlar-50. Moreover, the LSM method showed that the 4th quartile of *cis*-nonachlor was
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848 324 significantly increased compared to the 1st quartile *cis*-nonachlor. LSM analysis of the
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850 325 testosterone-androstenedione ratio showed decreasing trends in relation to *cis*-
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852 326 nonachlor, *trans*-nonachlor, Dieldrin, Mirex, and Parlar-50. Moreover, LSM analysis of
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854 327 the adrenal androgen-corticoid ratio showed an increasing trend of *cis*-nonachlor,
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856 328 Parlar-26, and Parlar-50, while SHBG showed a decreasing trend of β -HCH. Prolactin
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858 329 showed decreased trends of p,p' -DDE and o,p' -DDT. Statistically significant
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861 330 relationships between hormones and OCPs in the quartile models were not found among
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863 331 girls.
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868 333 **4. Discussion**
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870 334 In linear models, we found that relatively low levels of OCPs were inversely associated
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872 335 with testosterone, cortisol, cortisone, SHBG, and prolactin, but positively associated
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874 336 with DHEA in newborn boys after stratification by sex. Positive associations between
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876 337 OCPs and each of estradiol-testosterone and adrenal androgen-glucocorticoid, as well as
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878 338 an inverse association between OCPs and each of androstenedione-DHEA, testosterone-
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880 339 androstenedione, corticoid-cortisone, and the FSH-inhibin B ratio, were also observed

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890 340 among boys. In quartile models, testosterone was inversely associated with OCPs,
891 341 whereas DHEA was positively associated with them. The estradiol-testosterone and
892 342 adrenal androgen-glucocorticoid ratios tended to increase when OCP concentrations
893 343 were the higher quartile, while the testosterone-androstenedione ratio tended to
894 344 decrease. SHBG and prolactin showed inverse associations with OCPs. Among girls,
895 345 *p,p'*-DDD was inversely associated with DHEA levels and positively associated with
896 346 levels of cortisol; it was also inversely associated with the glucocorticoid-adrenal
897 347 androgen ratio. However, these associations were not observed in the quartile models,
898 348 which would have provided more credence to the findings of the linear models. Overall,
899 349 our data suggested that the natures of the associations between OCP and hormones
900 350 differ according to sex, and that clear associations were observed only among infant
901 351 boys.

902 352 The levels of OCPs in maternal blood have been measured in several studies
903 353 performed in various countries. In this study, the median *p,p'*-DDE value in maternal
904 354 whole blood was 619.26 pg/g-wet. In a study in Chiapas, Mexico, the median values of
905 355 DDT and DDE in maternal serum collected in 2002–2003 were 1.9 and 19.5 µg/L,
906 356 respectively (Longnecker et al. 2007). A study of Mexican-Americans in the US state of
907 357 California found that the maternal geometric mean values of *p,p'*-DDT, *o,p'*-DDT, and
908 358 *p,p'*-DDE were 22.0, 1.8, and 1436.9 ng/g-lipid, respectively (Eskenazi et al. 2006). A
909 359 study in China showed geometric means of *p,p'*-DDE, HCB, and β-HCH in maternal
910 360 serum, at 203.54 ng/g, 70.62 ng/g, and 67.67 ng/g, respectively (Guo et al. 2014). In
911 361 another recent study in South Africa, the median maternal *p,p'*-DDE level was 241.3
912 362 ng/g lipids (Bornman et al. 2016). A meta-analysis of 12 European cohorts found that
913 363 the median cord serum *p,p'*-DDE concentration was 527.9 ng/L (ranging from 49.8

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950 364 ng/L to 1208 ng/L in various studies) (Govarts et al. 2012). By using Govarts et al.'s
951 365 conversion factor to calculate cord serum levels from maternal whole blood levels in
952 366 our study (cord serum level = 0.36 × maternal whole blood level), the estimated
953 367 median level of *p,p'*-DDE in the cord serum of our study was 229.1 ng/L, which is
954 368 approximately half the median value of European cohorts. Taken together, the OCP
955 369 exposure levels in our cohort were relatively low in comparison.
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962 370 Only 1 study, performed in France, examined prenatal OCP exposure and
963 371 hormone levels at birth (Waremboeuf et al. 2016). The investigators found that higher
964 372 levels of HCB and HCE were associated with reduced levels of testosterone and
965 373 elevated levels of SHBG among boys. HCE was also associated with higher levels of
966 374 estradiol and a lower testosterone-estradiol ratio among girls. There was no association
967 375 between *p,p'*-DDE and any hormone levels. Their results were partly consistent with
968 376 ours. In our study, OCPs including HCB and HCE showed inverse associations with
969 377 testosterone; these were not significant except for Mirex among boys. Although we did
970 378 not find any association between estradiol and OCPs, their levels tracked together in a
971 379 manner also observed by Waremboeuf et al. (2016). Moreover, the positive association
972 380 between the estradiol-testosterone ratio and OCPs, which was statistically significant for
973 381 *p,p'*-DDE in this study, are consistent with the findings of Waremboeuf et al. (2016), as
974 382 is the significant decrease in the aromatase index (testosterone/estradiol) with increasing
975 383 HCE. On the other hand, SHBG had a significant inverse correlation with Mirex in our
976 384 study; which was in direct contrast to Waremboeuf et al.'s findings; we are unable to
977 385 explain this discrepancy, as the sampling period (2002–2006) and sample sizes (n=282
978 386 for their study and n=232 for ours) were comparable. Additionally, the exposure levels
979 387 of DDE, HCE, and HCB in our study and in that of Waremboeuf et al. were
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1009 388 comparable. However, the detection rates of HCB and HCE were higher in our study,
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1011 389 and Warembourg et al. did not measure chlordanes, Mirex, or toxaphenes. Because of
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1013 390 the low detection percentage, Warembourg et al. divided OCP levels into two or three
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1015 391 categories to examine the association of each with hormones; this could explain why
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1017 392 their results do not match ours more closely. We found no other studies that investigated
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1019 393 associations between OCPs and DHEA, androstenedione, cortisol, or cortisone. More
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1021 394 studies are therefore warranted to ascertain the association between OCPs and
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1023 395 steroidogenesis at birth.

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1026 396 In this study, DDT, DDE, and DDD were not associated with steroid hormones
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1028 397 except for *p,p'*-DDE, which showed a positive association with the estradiol-
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1030 398 testosterone ratio in boys. The increased estradiol-testosterone ratio suggested increased
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1032 399 CYP19A1 enzyme activity. A previous animal study showed that *p,p'*-DDE induces
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1034 400 CYP19A1 in hepatic microsomal samples of adult male rats (You et al. 2001), which is
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1036 401 consistent with our results. In girls, *p,p'*-DDD was inversely associated with DHEA and
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1038 402 the adrenal androgen-glucocorticoid ratio, but was positively associated with cortisone
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1040 403 levels. However, these associations were not observed in the quartile models; therefore,
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1042 404 false positive associations are likely. Longnecker et al. (2007) conducted a study in
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1044 405 Mexico where maternal DDT levels are relatively high, and found no evidence that in
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1046 406 utero exposure to DDE was related to anogenital distance or penile dimensions at birth
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1048 407 among male infants. Additionally, Bornman et al. (2016) conducted a similar study in
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1050 408 South Africa and reported no associations between *p,p'*-DDT/-DDE or *o,p'*-DDT and
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1052 409 anogenital distance measurements at birth in either boys or girls. Yet another study in
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1054 410 Denmark found no association between prenatal *p,p*-DDE levels and reproductive
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1056 411 hormones at approximately 20 years of age (Vested et al. 2014). Taken together, these

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1067 412 data indicate that DDTs may not alter infant steroid hormones at the levels detected in
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1069 413 these studies.

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1072 414 Among boys, *p,p'*-DDE and *o,p'*-DDT were inversely associated with prolactin
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1074 415 in quartile models. Although there have been no specific studies of the mechanisms of
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1076 416 OCP influence on prolactin, other investigations suggest that newborns with lower
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1078 417 levels of prolactin in their cord blood are more likely to have an increased risk of
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1080 418 respiratory distress syndrome than those with higher levels of prolactin (Parker et al.
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1082 419 1989; Padvi et al. 2017). Combined with data from the Canadian Health Measures
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1084 420 Survey that DDT exposure is inversely associated with lung function parameters in
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1086 421 adults (Ye et al. 2015), long term follow-up of children who were exposed to these
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1088 422 chemicals is needed to clarify the health-related consequences of such exposure.

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1091 423 Studies of OCPs other than DDTs are scarce. We found that *cis/trans*-
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1093 424 nonachlor, Dieldrin, Mirex, and toxaphenes (Parlar-26 and 50) were positively
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1095 425 associated with DHEA as well as with the adrenal androgen-glucocorticoid ratio, but
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1097 426 were inversely associated with testosterone, the testosterone-androstenedione ratio, and
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1099 427 prolactin in boys. Decreasing testosterone-androstenedione ratio suggests the
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1101 428 downregulation of HSB17B1; previous animal studies have shown that aldrin inhibits
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1103 429 HSB17B1 (Chatterjee et al. 1988), which is consistent with our findings. DHEA is the
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1105 430 main precursor of sex hormones and cortisol antagonists (Mastorakos and Ilias 2003).
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1107 431 The increasing adrenal androgen-glucocorticoid ratio suggests that chlordanes, Dieldrin,
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1109 432 Mirex, and toxaphenes may shift steroidogenesis towards androgenic hormones (C19-
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1111 433 steroids), although we observed no evidence of this. However, circulating hormones are
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1113 434 regulated by several cascade reactions. Translocation of cholesterol from the outer
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1115 435 mitochondrial membrane to the inner membranes is a critical step in steroidogenesis;

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1127 436 this process is enhanced by steroidogenic acute regulatory protein (StAR), which is
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1129 437 encoded by *STARD1*. Other key human steroidogenic genes are *CYP11A1*, *CYP17A1*,
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1131 438 *CYP11B1/CYP11B2*, *CYP21A2*, *SRD5A1*, and *SRD5A2*. Moreover, *SULT2A1* and
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1133 439 *SULT2B1* encode enzymes that convert DHEA to DHEA-sulfate and vice versa.
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1135 440 Although we have not measured methoxychlor, another synthetic organochlorine
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1137 441 insecticide and a derivative of DDT showed decreased expression of *CYP19A1*,
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1139 442 *HSD17B1*, *CYP11A1*-encoded cytochrome P450 family 17 subfamily A member 1,
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1141 443 *HSD3B1*, *CYP11A1*-encoded cytochrome P450 family 11 subfamily A member 1, and
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1143 444 StAR, but increased expression of *CYP1B1*-encoded cytochrome P450 family 1
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1145 445 subfamily B member 1 enzyme levels, *in vitro* (Basavarajappa et al. 2011; Vaithinathan
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1147 446 et al. 2008). Therefore, our overall findings do not rule out that many of the OCPs
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1149 447 measured in this study may alter steroid hormones by inhibiting such steroidogenesis
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1151 448 enzymes. Moreover, a previous study notably examined 15 organochlorines (OCs)
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1153 449 mixture similar to those found in the bladders of Arctic ringed seals. Exposing male rats
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1155 450 *in utero* to this OC mixture caused disruptions in the development of androgen-
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1157 451 dependent organs such as the testis, epididymis, seminal vesicle, and prostate (Anas et
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1159 452 al. 2005). They also investigated that OC mixture's direct inhibition of Leydig cell
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1161 453 steroidogenesis by disrupting cholesterol transport into the mitochondria via decreasing
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1163 454 StAR protein levels, and by converting cholesterol into pregnenolone by modulating
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1165 455 adrenodoxin reductase and *CYP11A1* proteins (Enangue Njembele et al. 2014). We
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1167 456 speculate that different OCPs influence steroidogenesis similarly. Additional studies on
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1169 457 altered steroidogenic and metabolic enzymes with different OCPs are required to test
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1171 458 these hypotheses.

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1185 459 In our linear regression model, HCH, Mirex, and toxaphenes were positively
1186 460 associated with FSH, as was the FSH-inhibin B ratio. Sertoli cells secrete inhibin B,
1187 461 which was first identified by its ability to negatively regulate FSH (Carlson et al. 2009).
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1189 462 Increasing levels of FSH together with the FSH-inhibin B ratio point to a negative
1190 463 feedback mechanism as well as the influence of OCPs on Sertoli cells. However, these
1191 464 associations were not statistically significant in our quartile model; hence, the effects of
1192 465 OCPs on Sertoli cells remain unclear.
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1200 466 Our study found a significant association between OCPs and hormones only
1201 467 among boys. Although the reason for this remains unclear, one possible explanation for
1202 468 this sex preference is that most OCPs have an affinity for hormone receptors including
1203 469 androgen receptor, estrogen receptor, and/or aryl hydrocarbon receptor (Mnif et al.
1204 470 2011; Kojima et al., 2004). Thus, OCPs may mimic or block the natural hormone's
1205 471 action (as an agonist or antagonist, respectively), and can interfere with the synthesis,
1206 472 transport, metabolism, and elimination of hormones (Mnif et al. 2011). Estrogen and
1207 473 androgen receptors are primarily involved in sexual differentiation and reproduction
1208 474 (Busillo et al. 2009). While estrogens and progestins are essential for normal female
1209 475 development, androgens are involved in various aspects of male reproductive
1210 476 physiology. One animal study showed that *p,p'*-DDE inhibits androgen binding to the
1211 477 AR, androgen-induced transcriptional activity, and androgen action in developing
1212 478 (Kelce W.R., et al., 1995). Androgen receptors are expressed in variety of tissues, but
1213 479 their levels differed in human fetal tissues (Wilson and McPhaul, 1996). During fetal
1214 480 genital development, the female phenotype is considered to be the baseline (or default)
1215 481 condition, and the development of maleness requires additional secretions produced by
1216 482 the testis (Carlson 2009). Thus, we speculate that the effect of anti-androgenic activities
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1245 483 of OCPs on males is more severe than the effect of their estrogenic activities on
1246 females. The physiological differences in the roles of these hormones may explain the
1247 observation of sex differences in OCP-hormone associations. One epidemiological
1248 study in Mexico found that higher exposure to *p,p'*-DDE *in utero* shortened the anal
1249 position index among boys but not among girls (Torres-Sanchez et al., 2008). The
1250 authors explained that this was due to the putative androgen deficiency occurring due to
1251 the reduction of transcriptional activity that occurs when AR is blocked by *p,p'*-DDE *in*
1252 *utero*. Although the exposure levels in Torres-Sanchez et al.'s study were 10 times
1253 higher than that in ours, our results were in line with those of their study. In vitro
1254 reporter gene assays showed that not only *p,p'*-DDE, but also other DDTs, dieldrin, and
1255 heptachlors showed AR antagonist effects (Kojima et al., 2004). Hence, we assume that
1256 those OCPs studied herein with AR antagonist properties share similar modes of action
1257 in terms of male-specific responses.

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1260 496 The strength of this study is that we examined a wide range of OCPs.
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1262 497 Investigating seven steroids enabled us to construct a clearer picture of the association
1263 between steroidogenesis and OCPs levels. Measuring steroid hormones by LC-MSMS
1264 is considered more accurate than other methods such as radioimmunoassays, which is
1265 another strength of this study. However, there are several limitations as well. First, the
1266 500 15 OCPs are moderately-to-highly correlated with each other (Spearman's rho: 0.230–
1267 0.918, $P < 0.001$). Thus, we were unable to clarify the effect of individual OCPs on
1268 hormones. Moreover, there may be other residual confounding factors. Previous studies
1269 suggested that reproductive hormones in cord blood may be affected by factors such as
1270 diurnal variation, duration of labor, placental weight, and the presence of pre-eclampsia
1271 (Hollier et al. 2014; Keelan et al. 2012). A recent cross-sectional study among adults
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1304 507 suggested an association between OCP concentrations and total cholesterol levels
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1306 508 (Arrebola et al. 2014), which may also modify steroid hormones levels. Second,
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1308 509 multiple tests of 15 OCPs and 13 hormones may have found statistically significant
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1310 510 associations by chance. However, the associations between OCPs and hormones
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1312 511 exhibited consistent trends, which suggest that the results are robust and that OCP
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1314 512 exposure is likely to alter infants' hormones in utero. Finally, there is a possibility of
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1316 513 selection bias in this study, as only participants with available cord blood samples were
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1318 514 included in the analysis. Because the life and the safety of the mother and child were of
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1320 515 utmost priority, cord blood samples were seldom acquired during Caesarian sections.
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1322 516 Thus, two infants were delivered Caesarian sections and others included in this study
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1324 517 were delivered vaginally; they had longer gestational ages and heavier birth weights
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1326 518 than the infants who were excluded, which indicated that the analysis was biased
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1328 519 towards healthier infants. Therefore, the effects of OCPs may have been underestimated
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1330 520 in this study.

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1334 522 **5. Conclusion**

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1336 523 We found that exposure to relatively low levels of OCPs such as *cis/trans*-nonachlor,
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1338 524 *p,p'*-DDE, *o,p'*-DDT, Dieldrin, β -HCH, Mirex, and toxaphenes *in utero* was
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1340 525 significantly associated with levels of hormones and their ratios in male fetuses. These
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1342 526 results suggest that OCPs that include but are not limited to DDTs ought to be
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1344 527 examined, as the existing data are limited. Disrupting the balance of steroid hormones
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1346 528 may cause adverse effects on reproductive growth, development, and other health
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1348 529 outcomes in later life. The clinical significance of these findings is unclear at present, as
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1350 530 it remains unknown whether these small hormonal alterations are of any future health

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1362 531 consequences in these individuals. Therefore, further studies investigating the long-term
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1364 532 effects of OCP exposure are required.

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1369 534 **Acknowledgements**

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1371 535 We would like to thank the mothers and children who participated in the study, as well
1372
1373 536 as all the staff at the Sapporo Toho Hospital. This study was supported in part by the
1374
1375 Japan Ministry of Health, Labour and Welfare; the Environment Research and
1376
1377 Technology Development Fund (5C-1252 and 5-1554); and Grants in Aid of Scientific
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1379 Research from the Japan Society for the Promotion of Science, the Ministry of
1380
1381 Education, Culture, Sports, Science, and Technology (13307015, 16209022, 19209024,
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1383 541 22249021, 26740028, and 26670321).

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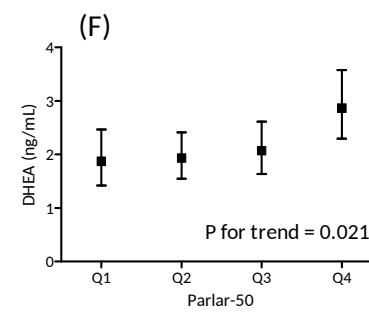
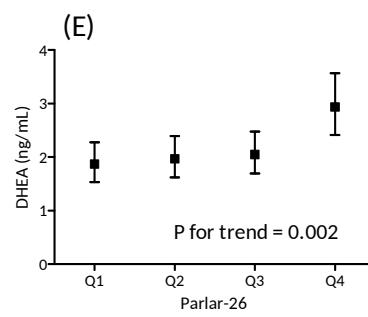
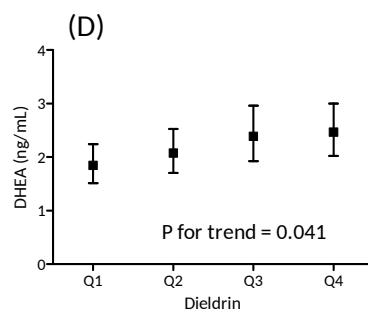
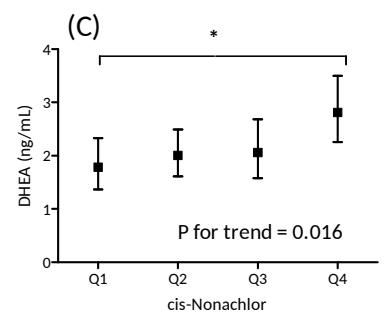
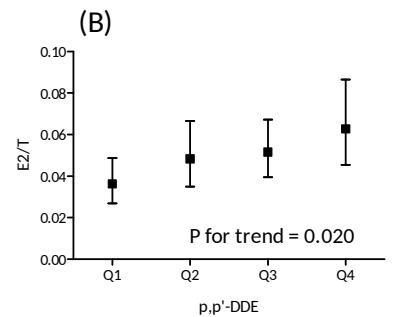
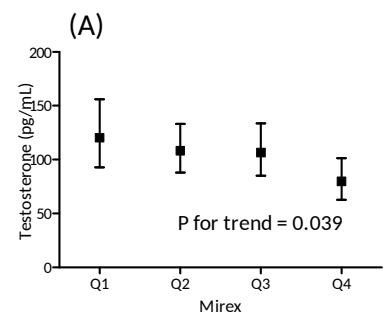
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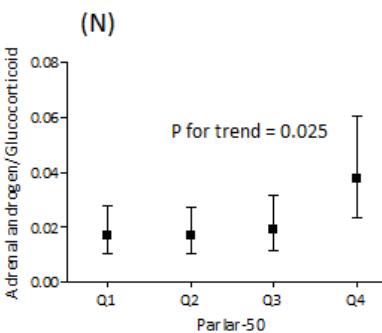
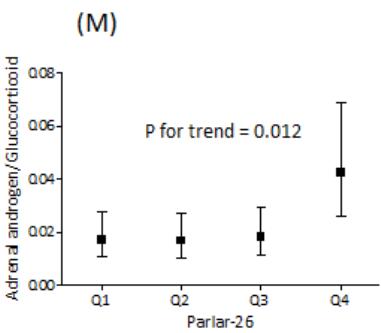
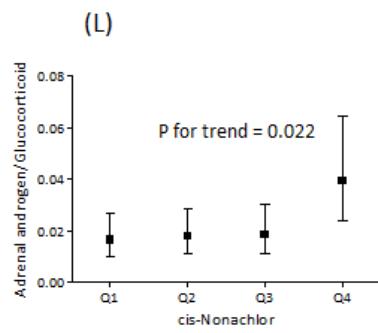
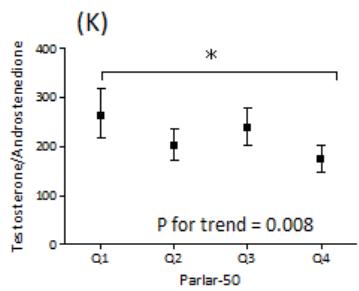
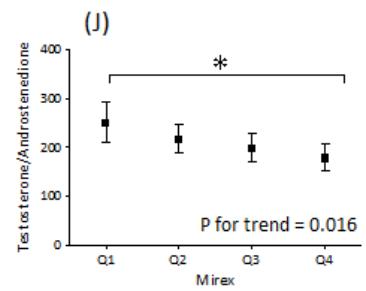
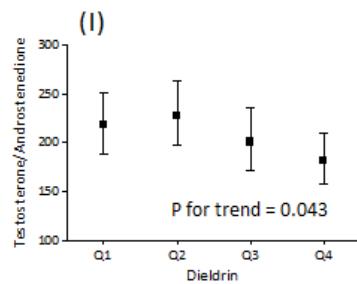
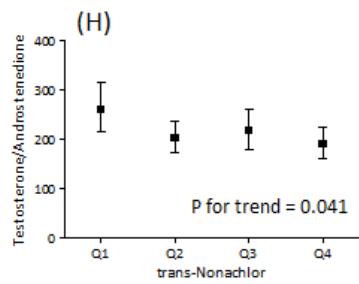
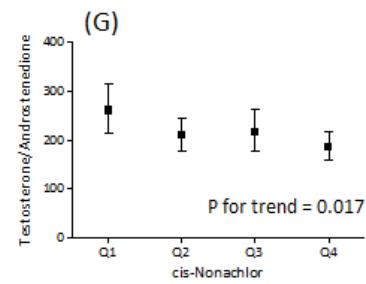
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1953 735 Figure legends
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1957 737 Fig. 1. Least square means (LSMs) of hormone levels according to maternal
1958 organochlorine pesticide (OCP) concentration quartiles in boys. The X-axes show the
1959 OCP quartiles, while the Y-axes show each hormone level calculated using the LSM in
1960 the boxes; the error bars are the 95% confidence intervals. The four OCP categories
1961 were, *cis*-nonachlor: Quartile 1 (Q1) (\leq 7.07 pg/g-wet), Q2 (7.08–10.37 pg/g-wet), Q3
1962 (10.38–15.06 pg/g-wet), and Q4 (\geq 15.07 pg/g-wet); *trans*-nonachlor: Q1 (\leq 52.25 pg/g-
1963 wet), Q2 (52.26–75.60 pg/g-wet), Q3 (52.26–75.60 pg/g-wet), and Q4 (\geq 110.29 pg/g-
1964 wet); *p,p'*-DDE: Q1 (\leq 0.99 pg/g-wet), Q2 (1.00–1.65 pg/g-wet), Q3 (1.66–2.54 pg/g-
1965 wet), and Q4 (\geq 2.55 pg/g-wet); *o,p'*-DDT: Q1 (\leq 2.28 pg/g-wet), Q2 (2.29–3.36 pg/g-
1966 wet), Q3 (3.37–4.66 pg/g-wet), and Q4 (\geq 4.67 pg/g-wet); Dieldrin: Q1 (\leq 12.17 pg/g-
1967 wet), Q2 (12.17–16.68 pg/g-wet), Q3 (16.69–22.05 pg/g-wet), and Q4 (\geq 22.06 pg/g-
1968 wet); β -hexachlorocyclohexane (β -HCH): Q1 (\leq 104.33 pg/g-wet), Q2 (104.34–154.31
1969 pg/g-wet), Q3 (154.32–238.06 pg/g-wet), and Q4 (\geq 238.07 pg/g-wet); Mirex: Q1 (\leq 4.12
1970 pg/g-wet), Q2 (4.13–6.04 pg/g-wet), Q3 (6.05–8.52 pg/g-wet), and Q4 (\geq 8.53 pg/g-
1971 wet); Parlar-26: Q1 (\leq 2.84 pg/g-wet), Q2 (2.85–4.46 pg/g-wet), Q3 (4.47–7.11 pg/g-
1972 wet), and Q4 (\geq 7.12 pg/g-wet); and Parlar-50: Q1 (\leq 4.31 pg/g-wet), Q2 (4.31–6.56
1973 pg/g-wet), Q3 (6.56–9.83 pg/g-wet), and Q4 (\geq 9.83 pg/g-wet). (A) Testosterone
1974 according to Mirex, (B) Estradiol-testosterone ratio (E2/T) according to *p,p'*-
1975 dichlorodiphenylchloroethylene (DDE), (C) dehydroepiandrosterone (DHEA)
1976 according to *cis*-nonachlor, (D) DHEA according to Dieldrin, (E) DHEA according to
1977 Parlar-26, (F) DHEA according to Parlar-50, (G) Testosterone-androstenedione ratio
1978 (T/Adione) according to *cis*-Nonachlor, (H) T/Adione according to *trans*-Nonachlor, (I)

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2012 759 T/Adione according to Dieldrin, (J) T/Adione according to Mirex, (K) T/Adione
2013 760 according to Parlar-50, (L) Adrenal androgen-glucocorticoid ratio according to *cis*-
2014 761 Nonachlor, (M) Adrenal androgen-glucocorticoid ratio according to Parlar-26, (N)
2015 762 Adrenal androgen-glucocorticoid-ratio according to Parlar-50, (O) sex hormone-binding
2016 763 globulin (SHBG) according to β -HCH, (P) Prolactin according to *p,p'*-DDE, (Q)
2017 764 Prolactin according to *o,p'*-DDT. The first quartile is compared to the second, third, and
2018 765 fourth quartile OCPs as calculated using the Dunnett-Hsu method; the statistical
2019 766 significance of the *P* value was **P*<0.017 based on Bonferroni's correction. LSMS were
2020 767 adjusted for maternal age, parity, and gestational age.
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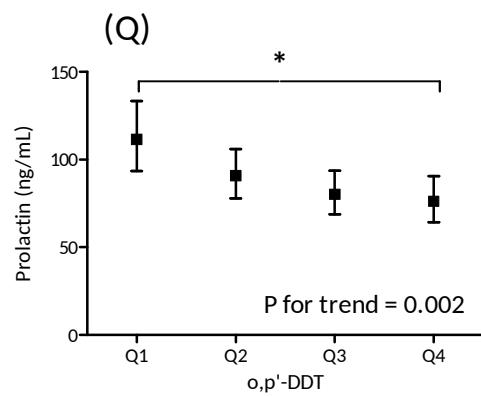
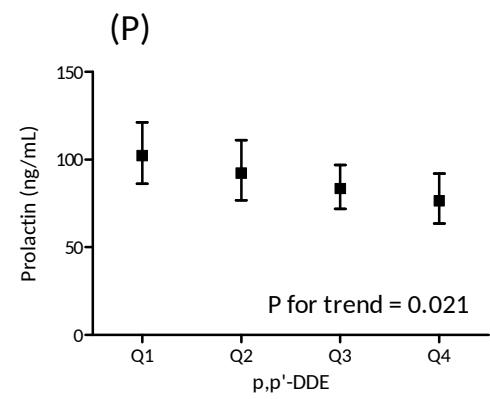
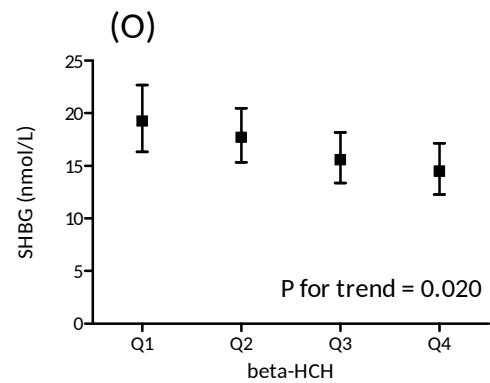


Fig. 1.

Table 1 Maternal and infant characteristics

	This study (n=232)			Original cohort (n=514)		
	No.	%	Mean ± SD	No.	%	Mean ± SD
Mother						
Age at delivery (years)	232		30.45 ± 4.81	510		30.4 ± 4.9
Pre-pregnancy BMI (kg/m ²)	232		21.03 ± 2.92	506		21.2 ± 3.2
Educational level (years)						
≤12	100	43.1		225	44.3	
>12	132	56.9		283	55.7	
Annual Household income (million yen per year)						
<5	166	71.6		345	68.3	
≥5	66	28.4		160	31.7	
Smoking during pregnancy						
No	190	81.9		404	79.7	
Yes	42	18.1		103	20.3	
Alcohol consumption during pregnancy						
No	154	66.4		351	69.1	
Yes	78	33.6		157	30.9	
Parity						
Primiparous	120	51.7		240	47.7	
Multiparous	112	48.3		263	52.2	
Type of delivery						
Vaginal	230	99.1		397	78.8	
Caesarian section	2	0.9		107	21.2	
Blood sampling timing						
During pregnancy	159	68.5		296	69.5	
After delivery	73	31.5		130	30.5	
Infant						
Sex						
Boys	106	45.7		246	48.1	
Girls	126	54.3		265	51.9	
Birth weight (grams)	232		3130.5 ± 332.5	511		3025.6 ± 420.7
Gestational Age (weeks)	232		39.3 ± 1.1	511		38.9 ± 1.5
cryptorchidism	1	0.9		1	0.4	
hypospadias	0	0.0		0	0.0	

Table 2 Concentrations of organochlorine pesticides in maternal blood

Persistent organochlorine pesticides	Detection limit (pg/g-wet)	>DL (%)	Minimum	Percentile			
				25th	50th	75th	Maximum
Aldrin	1.00	0.4				<DL	12.83
Chlordanes							
oxychlordane	0.90	100.0	7.93	28.87	40.04	57.32	250.94
<i>cis</i> -Chlordane	0.70	62.1		<DL	1.15	2.29	17.53
<i>trans</i> -Chlordane	0.50	49.6			<DL	0.84	3.79
<i>cis</i> -Nonachlor	0.40	100.0	1.63	7.07	10.37	15.07	37.58
<i>trans</i> -Nonachlor	0.50	100.0	13.45	52.09	75.60	110.54	513.52
DDTs							
<i>o,p'</i> -DDD	0.50	14.2				<DL	1.16
<i>p,p'</i> -DDD	0.40	88.8	<DL	0.98	1.65	2.54	9.04
<i>o,p'</i> -DDE	0.40	86.6	<DL	0.72	1.25	1.78	4.60
<i>p,p'</i> -DDE	0.60	100.0	99.52	409.79	619.26	968.05	2686.23
<i>o,p'</i> -DDT	0.60	96.6	<DL	2.28	3.36	4.67	17.15
<i>p,p'</i> -DDT	0.40	100.0	2.38	16.22	23.17	33.94	104.76
Dieldrin	0.80	100.0	4.11	12.16	16.68	22.21	71.52
Endrin	1.00	0.0					<DL
Heptachlors							
Heptachlor	0.80	0.9				<DL	1.14
<i>trans</i> -HCE	1.00	0.0					<DL
<i>cis</i> -HCE	0.40	100.0	6.17	18.81	26.25	37.45	200.53
HCB	0.90	100.0	34.94	83.04	103.99	131.61	245.48
HCHs							
α -HCH	0.70	68.5		<DL	0.91	1.31	3.10
β -HCH	0.60	100.0	19.95	104.25	154.31	238.45	717.67
γ -HCH	0.90	57.3		<DL	1.09	1.73	100.92
δ -HCH	0.70	1.3				<DL	1.11
Mirex	0.50	100.0	0.88	4.11	6.04	8.53	30.11
Toxaphenes							
Parlar-26	1.00	97.0	<DL	2.84	4.46	7.13	20.82
Parlar-41	0.70	28.4			<DL	0.73	1.96
Parlar-40	2.00	0.9				<DL	2.43
Parlar-44	2.00	2.2				<DL	2.77
Parlar-50	2.00	96.1	<DL	4.30	6.56	9.86	29.29
Parlar-62	6.00	0.0				<DL	

DL, detection limit; DDD, dichlorodiphenyl dichloroethane; DDE, dichlorodiphenyl dichloroethylene; DDT, dichlorodiphenyl trichloroethane; HCB, hexachlorobenzene; HCE, heptachlor epoxide; HCH, hexachlorocyclohexane.

Table 3 Distribution of steroids and reproductive hormones

	DL	n	>DL (%)	Total (n=232)				Boys (n=106)				Girls (n=126)				P-value		
				25%	Med	75%	n	>DL (%)	25%	Med	75%	n	>DL (%)	25%	Med	75%		
Steroid hormones																		
Progesterone (ng/mL)	0.01	232	100.0	177.381	219.036	278.019	106	100.0	184.087	224.121	280.633	126	100.0	169.619	216.065	276.129	0.460	
Testosterone (pg/mL)	0.01	232	100.0	59.750	82.050	114.025	106	100.0	76.788	101.875	132.050	126	100.0	50.975	68.675	92.400	<0.001	
Estradiol (ng/mL)	0.01	232	99.6	3.220	4.546	7.053	106	99.1	3.248	4.458	7.311	126	100.0	3.140	4.645	6.672	0.393	
DHEA (ng/mL)	0.01	232	100.0	1.763	2.160	2.970	106	100.0	1.590	2.090	2.830	126	100.0	1.898	2.270	3.140	0.040	
Androstenedione (ng/mL)	0.01	232	99.6	0.360	0.460	0.588	106	99.1	0.380	0.470	0.613	126	100.0	0.350	0.450	0.580	0.414	
Cortisol (ng/mL)	0.25	232	98.7	22.808	41.485	65.258	106	99.6	22.415	41.065	66.213	126	98.4	24.725	42.455	63.953	0.773	
Cortisone (ng/mL)	0.10	232	96.1	76.393	95.145	124.655	106	98.1	72.825	94.175	122.963	126	94.4	77.150	95.650	125.123	0.807	
Steroid Hormone Binding Globulin (nmol/L)	1.10	232	99.6	13.500	15.700	18.875	106	100.0	13.575	16.200	19.300	126	99.2	13.300	15.450	18.425	0.239	
Lutheeling Hormone (mIU/mL)	0.50	226	17.3	<DL				103	36.9	<DL				123	0.8	<DL		<0.001
Follicle Stimulating Hormone (mIU/mL)	0.50	225	22.2	<DL				103	48.5	<DL				122	0.0	<DL		<0.001
Inhibin B (pg/mL)	11	232	61.6	<DL	23.500	45.325	106	98.1	34.200	44.350	61.300	126	31.0	<DL		13.900	<0.001	
Insulin-like factor 3 (ng/mL)	0.01	119	100.0	0.230	0.280	0.330	103	100.0	0.250	0.290	0.340	16	100.0	0.180	0.185	0.235	<0.001	
Prolactin (ng/mL)	1.0	226	99.6	64.600	87.600	118.250	103	100.0	63.400	87.200	119.000	123	99.2	64.600	87.800	116.000	0.921	

P values were calculated by Mann-Whitney U test;

DHEA, dehidroepiandrosterenedione; DL, detection limit

Table 4 Associations between OCPs exposure and steroid and reproductive hormone levels among boys

	Progesterone			Estradiol			Testosterone			Estradiol/Testosterone			DEHA			Androstenedione			Androstenedione/DHEA							
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI						
Oxychlordane	-0.090	-0.416	0.236	0.007	-0.359	0.372	-0.176	-0.421	0.068	0.183	-0.140	0.505	0.243	0.028	0.459	*	0.006	-0.214	0.226	-0.238	-0.470	-0.005	*			
cis-Nonachlor	-0.149	-0.424	0.125	-0.005	-0.314	0.304	-0.095	-0.303	0.113	0.090	-0.184	0.363	0.247	0.066	0.427	**	0.070	-0.116	0.255	-0.177	-0.375	0.021	†			
trans-Nonachlor	-0.162	-0.436	0.112	-0.027	-0.335	0.282	-0.161	-0.367	0.045	0.134	-0.138	0.407	0.255	0.076	0.435	**	0.038	-0.147	0.224	-0.217	-0.413	-0.021	*			
p,p'-DDD	-0.097	-0.279	0.085	-0.151	-0.354	0.051	-0.065	-0.203	0.073	-0.086	-0.267	0.095	-0.023	-0.147	0.101	-0.016	-0.139	0.108	0.007	-0.126	0.141					
o,p'-DDE	-0.046	-0.234	0.143	0.050	-0.161	0.260	-0.029	-0.171	0.114	0.078	-0.108	0.265	0.080	-0.046	0.207	0.013	-0.114	0.140	-0.067	-0.204	0.069					
p,p'-DDE	0.034	-0.218	0.286	0.192	-0.088	0.472	-0.105	-0.295	0.084	0.297	0.054	0.541	*	0.108	-0.061	0.278	-0.041	-0.211	0.128	-0.150	-0.331	0.032				
o,p'-DDT	-0.056	-0.248	0.136	0.048	-0.167	0.263	-0.035	-0.180	0.110	0.083	-0.107	0.273	0.118	-0.010	0.246	†	0.010	-0.119	0.140	-0.108	-0.246	0.031				
p,p'-DDT	-0.077	-0.339	0.184	0.031	-0.262	0.324	-0.065	-0.263	0.132	0.096	-0.163	0.356	0.151	-0.024	0.325	†	0.016	-0.161	0.192	-0.135	-0.324	0.054				
Dieldrin	-0.178	-0.482	0.125	0.069	-0.273	0.410	-0.023	-0.254	0.208	0.092	-0.211	0.394	0.267	0.067	0.467	**	0.155	-0.048	0.359	-0.112	-0.333	0.109				
cis-HCE	-0.090	-0.416	0.236	0.007	-0.359	0.372	-0.176	-0.421	0.068	0.183	-0.140	0.505	0.243	0.028	0.459	*	0.006	-0.214	0.226	-0.238	-0.470	-0.005	*			
HCB	0.025	-0.438	0.489	0.114	-0.404	0.633	-0.139	-0.489	0.210	0.254	-0.203	0.711	0.311	0.004	0.619	*	0.058	-0.254	0.370	-0.253	-0.586	0.081				
β -HCH	0.118	-0.157	0.392	0.168	-0.139	0.474	-0.091	-0.299	0.117	0.259	-0.010	0.528	†	0.136	-0.049	0.320	-0.004	-0.190	0.181	-0.140	-0.339	0.059				
Mirex	-0.108	-0.418	0.203	-0.007	-0.355	0.342	-0.262	-0.492	-0.032	*	0.256	-0.050	0.561	0.213	0.007	0.420	*	-0.061	-0.270	0.149	-0.274	-0.494	-0.054	*		
Parlar-26	-0.124	-0.338	0.091	0.124	-0.116	0.364	0.002	-0.161	0.165	0.122	-0.090	0.335	0.235	0.097	0.373	**	0.100	-0.044	0.244	-0.135	-0.290	0.020	†			
Parlar-50	-0.099	-0.315	0.116	0.072	-0.170	0.313	-0.020	-0.183	0.144	0.091	-0.123	0.305	0.236	0.097	0.374	**	0.103	-0.041	0.248	-0.132	-0.288	0.023	†			
Testosterone/Androstenedione			Cortisol			Cortisone			Cortisone/Cortisol			Adrenal androgen/Glucocorticoid			SHBG			Prolactin								
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI						
Oxychlordane	-0.182	-0.339	-0.024	*	-0.363	-0.763	0.038	-0.398	-0.858	0.062	†	-0.035	-0.316	0.246	0.571	0.042	1.100	*	-0.082	-0.245	0.081	-0.077	-0.261	0.107		
cis-Nonachlor	-0.165	-0.297	-0.032	*	-0.218	-0.559	0.124	-0.310	-0.700	0.079	-0.092	-0.330	0.145	0.487	0.040	0.935	*	-0.040	-0.178	0.098	-0.150	-0.303	0.003	†		
trans-Nonachlor	-0.199	-0.330	-0.069	**	-0.372	-0.708	-0.036	*	-0.426	-0.810	-0.041	*	-0.054	-0.291	0.184	0.609	0.168	1.050	**	-0.027	-0.165	0.111	-0.075	-0.230	0.080	
p,p'-DDD	-0.049	-0.140	0.041	0.061	-0.166	0.289	0.014	-0.248	0.275	-0.048	-0.205	0.110	-0.049	-0.352	0.255	0.055	-0.036	0.146	0.021	-0.082	0.124					
o,p'-DDE	-0.042	-0.134	0.051	-0.108	-0.342	0.126	-0.148	-0.415	0.119	-0.040	-0.202	0.122	0.200	-0.109	0.510	-0.030	-0.124	0.064	-0.120	-0.224	-0.015	*				
p,p'-DDE	-0.064	-0.188	0.060	-0.290	-0.600	0.019	†	-0.249	-0.606	0.108	0.041	-0.176	0.258	0.354	-0.059	0.766	†	-0.080	-0.206	0.045	-0.226	-0.361	-0.092	**		
o,p'-DDT	-0.045	-0.140	0.049	-0.172	-0.410	0.065	-0.153	-0.426	0.120	0.019	-0.146	0.185	0.260	-0.055	0.574	-0.017	-0.114	0.079	-0.172	-0.275	-0.069	**				
p,p'-DDT	-0.081	-0.210	0.048	-0.073	-0.400	0.253	-0.192	-0.565	0.180	-0.119	-0.344	0.105	0.276	-0.155	0.707	-0.019	-0.150	0.112	-0.155	-0.299	-0.011	*				
Dieldrin	-0.178	-0.325	-0.032	*	0.028	-0.352	0.409	-0.207	-0.641	0.227	-0.235	-0.494	0.023	†	0.367	-0.134	0.867	-0.049	-0.202	0.103	-0.209	-0.377	-0.040	*		
cis-HCE	-0.182	-0.339	-0.024	*	-0.363	-0.763	0.038	†	-0.398	-0.858	0.062	†	-0.035	-0.316	0.246	0.571	0.042	1.100	*	-0.082	-0.245	0.081	-0.077	-0.261	0.107	
HCB	-0.198	-0.424	0.028	†	-0.255	-0.830	0.321	-0.185	-0.845	0.476	0.070	-0.329	0.469	0.474	-0.289	1.236	-0.167	-0.397	0.062	-0.239	-0.495	0.017	†			
β -HCH	-0.087	-0.222	0.048	-0.153	-0.495	0.189	-0.076	-0.469	0.317	0.077																

Table 5 Associations between OCPs exposure and steroid and reproductive hormone levels among girls

	Progesterone			Estradiol			Testosterone			Estradiol/Testosterone			DEHA			Androstenedione					
	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI			
Oxychlordane	0.136	-0.149	0.420	0.047	-0.175	0.270	0.050	-0.219	0.320	-0.003	-0.222	0.216	-0.087	-0.377	0.203	0.032	-0.189	0.254			
cis-Nonachlor	0.194	-0.069	0.456	0.019	-0.188	0.225	0.035	-0.215	0.284	-0.016	-0.220	0.187	-0.136	-0.404	0.132	0.006	-0.199	0.211			
trans-Nonachlor	0.192	-0.065	0.449	0.035	-0.167	0.237	0.065	-0.179	0.310	-0.030	-0.229	0.169	-0.163	-0.425	0.099	0.038	-0.163	0.239			
p,p'-DDD	0.099	-0.046	0.244	-0.059	-0.172	0.054	-0.041	-0.178	0.096	-0.018	-0.130	0.094	-0.160	-0.305	-0.014	*	-0.069	-0.181	0.043		
o,p'-DDE	-0.042	-0.212	0.127	-0.082	-0.214	0.049	-0.038	-0.198	0.122	-0.044	-0.174	0.086	0.000	-0.173	0.173	-0.066	-0.197	0.065			
p,p'-DDE	0.018	-0.201	0.237	-0.116	-0.285	0.054	0.035	-0.172	0.241	-0.151	-0.317	0.015	†	-0.156	-0.377	0.065	-0.077	-0.246	0.092		
o,p'-DDT	-0.071	-0.282	0.140	-0.109	-0.272	0.055	-0.024	-0.222	0.175	-0.085	-0.246	0.076	0.013	-0.202	0.227	-0.072	-0.235	0.091			
p,p'-DDT	0.042	-0.205	0.289	-0.063	-0.256	0.129	0.075	-0.158	0.308	-0.138	-0.327	0.050	-0.102	-0.353	0.149	-0.041	-0.233	0.150			
Dieldrin	-0.012	-0.310	0.287	-0.028	-0.261	0.204	-0.092	-0.373	0.188	0.064	-0.165	0.293	0.068	-0.235	0.371	-0.204	-0.432	0.024			
cis-HCE	0.136	-0.149	0.420	0.047	-0.175	0.270	0.050	-0.219	0.320	-0.003	-0.222	0.216	-0.087	-0.377	0.203	0.032	-0.189	0.254			
HCB	0.216	-0.174	0.606	-0.052	-0.357	0.253	0.026	-0.344	0.395	-0.077	-0.378	0.223	-0.052	-0.450	0.346	0.051	-0.252	0.354			
β -HCH	-0.032	-0.286	0.222	-0.050	-0.248	0.147	0.057	-0.182	0.296	-0.107	-0.301	0.087	-0.016	-0.274	0.242	-0.006	-0.203	0.190			
Mirex	0.218	-0.068	0.503	0.043	-0.182	0.267	-0.030	-0.302	0.241	0.073	-0.148	0.294	-0.166	-0.457	0.126	0.025	-0.198	0.248			
Parlar-26	0.031	-0.167	0.229	-0.006	-0.160	0.149	-0.012	-0.198	0.175	0.006	-0.146	0.158	0.022	-0.180	0.223	-0.029	-0.182	0.124			
Parlar-50	0.046	-0.177	0.270	-0.012	-0.186	0.162	-0.002	-0.213	0.209	-0.010	-0.182	0.162	0.021	-0.207	0.248	-0.015	-0.189	0.158			
Androstenedione/DHEA			Testosterone/Androstenedione			Cortisol			Cortisone			Cortisone/Cortisol			Adrenal androgen/Glucocorticoid						
	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI	β	95%CI			
Oxychlordane	0.119	-0.217	0.455	0.018	-0.140	0.175	0.214	-0.265	0.692	0.374	-0.302	1.050	0.160	-0.145	0.465	-0.361	-1.139	0.416			
cis-Nonachlor	0.142	-0.169	0.453	0.029	-0.117	0.175	0.229	-0.215	0.672	0.463	-0.161	1.087	0.234	-0.046	0.515	-0.463	-1.181	0.255			
trans-Nonachlor	0.201	-0.103	0.505	0.027	-0.116	0.170	0.276	-0.157	0.709	0.517	-0.092	1.126	†	0.241	-0.033	0.516	†	-0.533	-1.234	0.168	
p,p'-DDD	0.091	-0.080	0.262	0.028	-0.052	0.108	0.208	-0.034	0.450	†	0.358	0.017	0.699	*	0.150	-0.004	0.304	†	-0.412	-0.803	-0.021
o,p'-DDE	-0.066	-0.266	0.134	0.027	-0.066	0.121	-0.176	-0.460	0.108	-0.030	-0.434	0.374	0.146	-0.034	0.327	0.088	-0.376	0.551			
p,p'-DDE	0.079	-0.179	0.337	0.112	-0.007	0.231	†	0.106	-0.261	0.474	0.285	-0.233	0.803	0.179	-0.054	0.411	-0.333	-0.928	0.261		
o,p'-DDT	-0.084	-0.333	0.164	0.048	-0.068	0.164	-0.269	-0.620	0.083	-0.058	-0.560	0.443	0.210	-0.013	0.433	†	0.143	-0.432	0.718		
p,p'-DDT	0.061	-0.231	0.353	0.116	-0.019	0.251	†	-0.045	-0.461	0.371	0.207	-0.380	0.795	0.253	-0.009	0.514	†	-0.167	-0.842	0.508	
Dieldrin	-0.272	-0.621	0.076	0.112	-0.051	0.275	-0.203	-0.704	0.297	-0.274	-0.982	0.434	-0.071	-0.391	0.249	0.256	-0.557	1.069			
cis-HCE	0.119	-0.217	0.455	0.018	-0.140	0.175	0.214	-0.265	0.692	0.374	-0.302	1.050	0.160	-0.145	0.465	-0.361	-1.139	0.416			
HCB	0.102	-0.359	0.564	-0.025	-0.241	0.191	0.317	-0.339	0.973	0.499	-0.428	1.425	0.182	-0.237	0.600	-0.420	-1.486	0.647			
β -HCH	0.010	-0.290	0.309	0.063	-0.077	0.202	0.004	-0.422	0.431	0.012	-0.592	0.615	0.007	-0.265	0.280	-0.014	-0.707	0.679			
Mirex	0.191	-0.148	0.529	-0.055	-0.214	0.103	0.229	-0.254	0.712	0.510	-0.170	1.189	0.280	-0.025	0.585	†	-0.525	-1.306	0.257		
Parlar-26	-0.051	-0.285	0.183	0.017	-0.092	0.127	0.044	-0.289	0.377	0.091	-0.380	0.562	0.047	-0.165	0.260	-0.040	-0.581	0.501			
Parlar-50	-0.036	-0.300	0.228	0.014	-0.110	0.137	0.028	-0.348	0.404	0.123	-0.408	0.655	0.095	-0.144	0.334	-0.050	-0.661	0.560			
	SHBG			Prolactin																	
	β	95%CI	β	95%CI	β	95%CI															
Oxychlordane	0.060	-0.118	0.238	0.092	-0.149	0.332															
cis-Nonachlor	0.057	-0.108	0.221	0.076	-0.147	0.298															
trans-Nonachlor	0.059	-0.102	0.220	0.085	-0.1																

* $P < 0.05$, ** $P < 0.01$, † $P < 0.1$

Supplemental Table S1 Concentrations of organochlorine pesticides in maternal blood all measured of original cohort (n=379)

Persistent organochlorine pesticides	Detection limit (pg/g-wet)	>DL (%)	Minimum	Percentile			
				25th	50th	75th	Maximum
Aldrin	1.00	0.3				<DL	12.83
Chlordanes							
oxychlordane	0.90	100.0	7.93	27.05	39.67	56.02	250.94
<i>cis</i> -Chlordane	0.70	58.8		<DL	1.15	2.29	17.53
<i>trans</i> -Chlordane	0.50	45.4			<DL	0.71	3.79
<i>cis</i> -Nonachlor	0.40	100.0	1.63	6.76	9.97	14.36	38.07
<i>trans</i> -Nonachlor	0.50	100.0	13.14	49.70	71.52	107.59	513.52
DDTs							
<i>o,p'</i> -DDD	0.50	12.4				<DL	1.16
<i>p,p'</i> -DDD	0.40	89.7	<DL	0.94	1.48	2.29	9.04
<i>o,p'</i> -DDE	0.40	85.0	<DL	0.75	1.27	1.82	6.20
<i>p,p'</i> -DDE	0.60	100.0	99.52	401.53	650.99	1011.48	4575.67
<i>o,p'</i> -DDT	0.60	7.6	<DL	2.27	3.48	4.86	17.15
<i>p,p'</i> -DDT	0.40	100.0	2.38	16.63	23.16	33.99	121.52
Dieldrin	0.80	100.0	4.11	12.08	16.42	22.62	71.52
Endrin	1.00	0.0				<DL	
Heptachlors							
Heptachlor	0.80	0.5				<DL	1.14
<i>trans</i> -HCE	1.00	0.0				<DL	
<i>cis</i> -HCE	0.40	100.0	6.17	18.78	26.44	37.28	200.53
HCB	0.90	100.0	34.94	80.24	101.65	130.06	245.48
HCHs							
α -HCH	0.70	69.1		<DL	0.90	1.32	3.89
β -HCH	0.60	100.0	19.95	105.05	154.45	244.76	1667.12
γ -HCH	0.90	59.1		<DL	1.05	1.63	100.92
δ -HCH	0.70	0.8				<DL	1.11
Mirex	0.50	100.0	0.88	4.07	5.95	8.26	34.97
Toxaphenes							
Parlar-26	1.00	97.1	<DL	2.87	4.39	6.65	20.82
Parlar-41	0.70	26.9			<DL	0.72	1.96
Parlar-40	2.00	0.5				<DL	2.43
Parlar-44	2.00	1.6				<DL	2.84
Parlar-50	2.00	96.0	<DL	4.36	6.52	9.68	29.29
Parlar-62	6.00	0.0				<DL	

DL, detection limit; DDD, dichlorodiphenyl dichloroethane; DDE, dichlorodiphenyl dichloroethylene; DDT, dichlorodiphenyl trichloroethane; HCB, hexachlorobenzene; HCE, heptachlor epoxide; HCH, hexachlorocyclohexane.

Supplemental table S2 Maternal and infant characteristics and concentrations of OCPs

characteristics	n	o,p'-DDT	p,p'-DDT	o,p'-DDE	p,p'-DDE	p,p'-DDD	cis-Nonachlor	trans-Nonachlor	oxychlordane
Mother		p 0.009	p 0.102	p 0.030	p 0.270**	p 0.137*	p 0.267**	p 0.311**	p 0.350**
Age at delivery (years)		0.062	0.118	0.119	0.003	0.052	0.089	-0.011	-0.093
Pre-pregnancy BMI (kg/m ²)		Med (min-max) 3.01 (2.09, 4.47)	Med (min-max) 21.59 (15.41, 32.84)	Med (min-max) 1.16 (0.70, 1.71)	Med (min-max) 565.05 (376.81, 942.71)	Med (min-max) 1.62 (0.88, 2.44)	Med (min-max) 9.88 (7.22, 14.64)	Med (min-max) 72.73 (52.69, 108.21)	Med (min-max) 39.83 (27.1, 56.02)
Educational level	≤12 years	100	3.01 (2.09, 4.47)	21.59 (15.41, 32.84)	1.16 (0.70, 1.71)	565.05 (376.81, 942.71)	1.62 (0.88, 2.44)	9.88 (7.22, 14.64)	72.73 (52.69, 108.21)
	>12 years	132	3.54 (2.42, 4.99)	25.32 (16.37, 34.77)	1.33 (0.96, 1.86)	634.71 (422.04, 976.14)	1.69 (1.03, 2.58)	10.78 (6.85, 15.13)	78.85 (51.80, 118.08)
Annual Household income	<5 million yen per year	166	3.27 (2.22, 4.66)	22.35 (16.38, 33.65)	1.26 (0.71, 1.74)	619.26 (411.26, 942.60)	1.62 (0.99, .53)	9.93 (6.71, 14.35)	*
	≥5 million yen per year	66	3.64 (2.41, 5.06)	24.18 (15.54, 38.90)	1.23 (0.77, 2.05)	633.52 (392.24, 1302.85)	1.68 (0.98, 2.58)	11.91 (8.21, 15.98)	70.06 (50.04, 111.09)
Smoking during pregnancy	No	190	3.46 (2.37, 4.66)	23.74 (16.38, 33.93)	1.26 (0.73, 1.77)	617.59 8409.71, 971.83)	1.61 (0.98, 2.57)	10.70 (7.17, 15.02)	76.63 (54.06, 112.12)
	Yes	42	3.03 (1.79, 5.03)	21.09 (14.17, 34.22)	1.17 (0.70, 1.84)	625.95 (409.34, 944.10)	1.79 (1.07, 2.39)	8.94 (6.01, 15.10)	69.19 (48.19, 109.46)
Alcohol consumption during pregnancy	No	154	3.10 (2.21, 4.67)	* 22.62 (15.54, 33.44)	1.22 (0.69, 1.72)	605.48 (412.66, 971.83)	1.62 (0.99, 2.55)	10.14 (6.90, 14.76)	74.14 (51.75, 112.12)
	Yes	78	3.86 (2.56, 4.98)	24.14 (17.27, 33.98)	1.27 (0.90, 1.87)	648.50 (406.30, 976.56)	1.67 (0.87, 2.41)	11.04 (7.29, 15.25)	78.85 (52.92, 108.95)
Parity	0	120	3.39 (2.38, 4.67)	24.38 (17.17, 33.92)	1.35 (0.81, 1.83)	652.69 (453.78, 1002.45)	1.64 (1.00, 2.52)	11.52 (7.50, 15.13)	87.74 (56.35, 117.19)
	≥1	112	3.24 (2.22, 4.78)	21.98 (14.92, 34.69)	1.13 (0.69, 1.71)	564.40 (336.51, 938.47)	1.65 (0.91, 2.62)	9.5 (6.37, 14.06)	*
Blood sampling period	During pregnancy	159	3.42 (2.25, 4.72)	22.49 (15.57, 33.95)	1.22 (0.73, 1.83)	626.50 (412.96, 941.47)	1.51 (0.99, 2.28)	10.01 (7.06, 14.87)	70.38 (21.93, 105.61)
	After delivery	73	3.36 (2.28, 4.48)	23.77 (19.33, 34.17)	1.35 (0.71, 1.73)	614.51 (339.28, 1008.37)	1.76 (0.87, 2.84)	11.15 (6.86, 16.79)	87.82 (51.95, 124.70)
Type of delivery	Vaginal	230	3.36 (2.27, 4.68)	22.84 816.14, 33.93)	1.25 (0.72, 1.77)	619.26 8410.14, 964.80)	1.62 (0.98, 2.53)	10.33 (7.03, 15.04)	75.08 (51.89, 110.01)
	Caesarian section	2	3.70 (3.03, 4.36)	36.58 (23.77, 49.38)	1.40 (0.91, 1.88)	688.72 (304.51, 1072.92)	2.87 (2.23, 3.5)	18.84 (12.12, 25.55)	119.01 (91.04, 146.97)
									55.41 (40.2, 70.62)
Infant		p 0.022	p -0.025	p 0.005	p 0.002	p -0.087	p -0.057	p -0.057	p -0.061
Birth weight		0.066	0.037	0.039	0.117	-0.013	0.025	0.035	0.045
Gestational Age		Med (min-max) 3.39 (2.42, 4.42)	Med (min-max) 23.90 (16.56, 33.88)	Med (min-max) 1.18 (0.72, 1.66)	Med (min-max) 641.93 (405.32, 934.78)	Med (min-max) 1.70 (1.04, 2.35)	Med (min-max) 10.29 (7.11, 15.11)	Med (min-max) 75.82 (53.44, 118.34)	Med (min-max) 40.22 (29.26, 57.54)
Sex	Male	106	3.36 (2.20, 4.97)	21.96 (15.78, 34.45)	1.30 (0.72, 1.87)	611.15 (4101.00, 998.41)	1.60 (0.93, 2.66)	10.42 (6.80, 15.04)	74.66 851.75, 107.06)
	Female								39.87 (28.34, 57.15)

Supplemental table S1 Maternal and infant characteristics and concentrations of OCPs (continued)

Characteristics	n	Dieldrin	cis-HCE	HCB	β-HCH	Mirex	Parlar-26	Parlar-50
Mother		p 0.108	p 0.254**	p 0.119'	p 0.464**	p 0.513**	p 0.176**	p 0.181**
Age at delivery (years)		0.221**	0.177**	0.101	0.147*	-0.074	0.240**	0.240**
Pre-pregnancy BMI (kg/m ²)		Med (min-max) 16.55 (12.52, 22.84)	Med (min-max) 26.07 (18.56, 40.28)	Med (min-max) 101.71 (80.26, 130.96)	Med (min-max) 154.79 (96.08, 262.30)	Med (min-max) 5.99 (4.08, 8.34)	Med (min-max) 4.85 (3.17, 7.55)	Med (min-max) 7.02 (4.61, 1028)
Educational level	≤12 years	100	16.55 (12.52, 22.84)	26.07 (18.56, 40.28)	101.71 (80.26, 130.96)	154.79 (96.08, 262.30)	5.99 (4.08, 8.34)	7.02 (4.61, 1028)
	>12 years	132	16.80 (12.11, 21.59)	26.91 (18.96, 37.13)	106.79 (85.26, 167.25)	154.31 (109.21, 220.69)	6.11 (4.31, 8.59)	4.05 (2.56, 6.68)
Annual Household income	<5 million yen per year	166	15.93 (12.02, 20.99)	26.07 (18.44, 35.00)	102.17 (80.36, 129.52)	153.64 (100.87, 209.23)	5.76 (3.88, 7.79)	** 4.32 (2.69, 6.82)
	≥5 million yen per year	66	17.89 (12.60, 24.41)	26.55 (19.17, 43.07)	111.51 (85.53, 145.40)	165.66 (113.97, 279.53)	7.44 (4.82, 11.64)	5.22 (3.14, 7.61)
Smoking during pregnancy	No	190	16.86 (12.35, 22.47)	26.79 (19.14, 37.90)	103.82 (83.69, 131.84)	158.46 (103.79, 241.30)	6.16 (4.51, 8.61)	4.50 (2.84, 6.60)
	Yes	42	15.51 (11.41, 21.31)	24.39 (16.80, 35.42)	104.48 (74.82, 132.31)	150.87 (103.84, 184.91)	5.69 (3.73, 7.82)	4.23 (2.73, 8.24)
Alcohol consumption during pregnancy	No	154	16.80 (12.14, 22.82)	27.18 (18.91, 39.08)	105.46 (81.20, 133.02)	162.80 (104.40, 244.79)	6.04 (4.11, 8.38)	4.19 (2.82, 6.82)
	Yes	78	16.35 (12.42, 21.66)	24.82 (18.42, 35.69)	99.34 (86.16, 130.72)	146.89 (102.56, 202.87)	6.09 (4.11, 8.61)	4.54 (2.91, 7.58)
Parity	0	120	17.16 (12.52, 21.71)	26.79 (19.16, 36.85)	109.93 (91.03, 131.94)	** 165.19 (114.52, 271.53)	* 5.98 (4.14, 8.81)	4.23 (2.73, 8.24)
	≥1	112	16.35 (12.05, 22.66)	25.17 (18.19, 40.63)	95.36 (72.71, 129.15)	143.66 (90.00, 211.57)	6.19 (4.11, 8.47)	4.02 (2.72, 6.68)
Blood sampling period	During pregnancy	159	17.59 (12.39, 22.87)	26.44 (18.91, 38.12)	104.06 (85.45, 130.56)	154.13 (108.08, 218.60)	5.88 (4.07, 7.86)	4.50 (2.93, 7.08)
	After delivery	73	15.51 (12.06, 21.32)	25.10 (18.50, 36.78)	103.73 (75.77, 138.73)	158.62 (94.33, 254.44)	6.76 (4.54, 10.19)	4.33 (2.56, 7.49)
Type of delivery	Vaginal	230	16.61 (12.16, 22.44)	26.15 (18.76, 37.42)	103.99 (82.85, 131.26)	154.31 (104.40, 237.67)	6.04 (4.11, 8.51)	4.41 (2.84, 7.09)
	Caesarian section	2	18.66 (16.82, 20.50)	36.20 (28.26, 44.14)	134.17 (85.55, 182.79)	154.45 (68.50, 240.39)	13.99 (4.84, 23.14)	7.67 (5.34, 10.00)
								10.15 (7.63, 12.67)
Infant		p -0.049	p -0.101	p -0.066	p -0.128	p -0.030	p -0.024	p -0.022
Birth weight		-0.035	-0.082	0.052	0.061	0.027	0.024	-0.018
Gestational Age		Med (min-max) 15.92 (12.07, 22.55)	Med (min-max) 26.39 (19.22, 40.80)	Med (min-max) 101.76 (85.53, 130.85)	Med (min-max) 153.65 (107.63, 237.67)	Med (min-max) 6.04 (4.53, 8.94)	Med (min-max) 4.50 (2.91, 6.86)	Med (min-max) 6.64 (4.33, 10.27)
Sex	Male</							

Supplemental Table S3 Distribution of steroids and reproductive hormones among all measured of original cohort (n=295)

	DL	n	>DL (%)	Total (n=295)			n	>DL (%)	Boys (n=106)			n	>DL (%)	Girls (n=126)			P-value
				25%	Med	75%			25%	Med	75%			25%	Med	75%	
Steroid hormones																	
Progesterone (ng/mL)	0.01	295	100	175.5	218.8	278.8	135	100	183.8	225.9	285.6	160	100	167.6	209.0	275.6	0.184
Testosterone (pg/mL)	0.01	295	100	59.8	83.9	114.1	135	100	76.5	98.9	126.3	160	100	52.1	70.1	96.8	<0.001
Estradiol (ng/mL)	0.01	295	99.7	3.29	4.70	7.10	135	99.3	3.33	4.86	7.42	160	100	3.16	4.68	6.61	0.227
DHEA (ng/mL)	0.01	295	100	1.77	2.19	2.99	135	100	1.59	2.08	2.76	160	100	1.91	2.34	3.22	
Androstenedione (ng/mL)	0.01	295	99.3	0.36	0.46	0.58	135	98.5	0.38	0.47	0.61	160	99.4	0.35	0.45	0.58	
Cortisol (ng/mL)	0.25	295	97.6	22.7	38.9	63.6	135	98.5	22.5	38.3	65.3	160	96.9	22.8	39.2	62.8	
Cortisone (ng/mL)	0.10	295	94.9	69.9	93.9	123.0	135	97.0	70.5	95.3	123.4	160	93.1	69.7	93.0	123.0	
Steroid Hormone Binding Globulin (nmol/L)	1.10	295	100	13.3	15.8	19.0	135	100	13.9	16.5	19.3	160	100	13.0	15.5	18.7	0.079
Luteinizing Hormone (mIU/mL)	0.50	288	16.3			<DL	132	34.8		<DL	0.84	156	0.6		<DL	<0.001	
Follicle Stimulating Hormone (mIU/mL)	0.50	287	20.9			<DL	132	45.5		<DL	0.66	155	0		<DL	<0.001	
Inhibin B (pg/mL)	11	295	59.7	5.5	23.2	44.6	135	98.5	33.9	44.0	58.3	160	26.9		<DL	12.4	<0.001
Insulin-like factor 3 (ng/mL)	0.01	157	100	0.23	0.27	0.32	132	100	0.25	0.29	0.34	25	100	0.17	0.18	0.23	<0.001
Prolactin (ng/mL)	1.0	289	99.7	63.1	85.8	116.0	132	100	65.4	85.2	115.0	157	99.6	61.4	86.0	118.0	0.986

P values were calculated by Mann-Whitney U test;

DHEA, dehydroepiandrosterone; DL, detection limit

Supplemental Table S4 Associations between OCPs exposure and steroid and reproductive hormone levels and sex interaction

	Progesterone				Estradiol				Testosterone				Estradiol/Testosterone				DEHA				Androstenedione				
	β	95%CI	$p^a)$	$p^b)$	β	95%CI	$p^a)$	$p^b)$	β	95%CI	$p^a)$	$p^b)$	β	95%CI	$p^a)$	$p^b)$	β	95%CI	$p^a)$	$p^b)$	β	95%CI	$p^a)$	$p^b)$	
Oxychlordane	0.024	-0.189	0.238		0.025	-0.178	0.229		-0.065	-0.249	0.118		0.091	-0.099	0.281		0.067	-0.119	0.253		0.019	-0.137	0.175		
cis-Nonachlor	0.022	-0.166	0.210	*	0.010	-0.170	0.190		-0.030	-0.192	0.132		0.040	-0.128	0.208		0.055	-0.108	0.218	*	0.039	-0.098	0.176		
trans-Nonachlor	0.014	-0.172	0.200	*	0.005	-0.173	0.184		-0.049	-0.209	0.112		0.054	-0.112	0.221		0.046	-0.116	0.208	*	0.038	-0.098	0.175		
p,p'-DDD	0.004	-0.111	0.119	†	-0.102	-0.211	0.008	†	-0.054	-0.153	0.045		-0.047	-0.150	0.055		-0.091	-0.191	0.009	†	-0.043	-0.126	0.041		
o,p'-DDE	-0.044	-0.169	0.082		-0.017	-0.136	0.102		-0.034	-0.142	0.074		0.017	-0.094	0.128		0.039	-0.070	0.149		-0.027	-0.118	0.064		
p,p'-DDE	0.017	-0.149	0.182		0.027	-0.130	0.183		-0.031	-0.172	0.110		0.058	-0.087	0.202	*	-0.029	-0.172	0.114		-0.060	-0.179	0.060		
o,p'-DDT	-0.069	-0.210	0.073		-0.035	-0.169	0.099		-0.024	-0.146	0.097		-0.011	-0.136	0.114		0.067	-0.056	0.190		-0.031	-0.134	0.071		
p,p'-DDT	-0.020	-0.198	0.158		-0.016	-0.186	0.154		0.007	-0.146	0.160		-0.023	-0.181	0.136		0.024	-0.130	0.179		-0.012	-0.141	0.118		
Dieldrin	-0.093	-0.304	0.118		0.022	-0.179	0.224		-0.059	-0.240	0.122		0.081	-0.106	0.269		0.164	-0.019	0.347	†	-0.022	-0.174	0.130	*	
cis-HCE	-0.040	-0.238	0.158	†	-0.009	-0.198	0.181		-0.053	-0.224	0.118		0.044	-0.133	0.221		0.098	-0.074	0.271		-0.028	-0.173	0.117		
HCB	0.122	-0.178	0.422		0.033	-0.253	0.319		-0.060	-0.317	0.198		0.093	-0.173	0.359		0.126	-0.135	0.387		0.053	-0.165	0.271		
β -HCH	0.035	-0.151	0.220		0.050	-0.126	0.227		-0.014	-0.173	0.145		0.065	-0.099	0.229		0.056	-0.106	0.217		-0.004	-0.139	0.130		
Mirex	0.061	-0.146	0.268	†	0.027	-0.171	0.226		-0.154	-0.332	0.024	†	0.182	-0.003	0.366	†	0.018	-0.163	0.200		-0.019	-0.171	0.133		
Parlar-26	-0.046	-0.190	0.099		0.061	-0.077	0.198		-0.004	-0.128	0.121		0.065	-0.064	0.193		0.127	0.003	0.252	*	0.036	-0.069	0.140		
Parlar-50	-0.027	-0.181	0.127		0.032	-0.114	0.179		-0.010	-0.142	0.123		0.042	-0.095	0.179		0.127	-0.006	0.260	†	0.044	-0.067	0.156		
Androstenedione/DHEA			Testosterone/Androstenedione				Cortisol				Cortisone				Cortisone/Cortisol				Adrenal androgen/Glucocorticoid						
Oxychlordane	-0.051	-0.263	0.161		-0.083	-0.195	0.029		-0.084	-0.402	0.234		-0.014	-0.436	0.408	†	0.071	-0.138	0.281		0.100	-0.386	0.585	†	
cis-Nonachlor	-0.018	-0.205	0.168		-0.067	-0.166	0.032		†	-0.009	-0.290	0.272		0.066	-0.306	0.438	*	0.076	-0.108	0.260	†	0.022	-0.405	0.450	*
trans-Nonachlor	-0.009	-0.194	0.175	*	-0.086	-0.184	0.011	†	*	-0.057	-0.334	0.220	*	0.040	-0.328	0.407	*	0.097	-0.085	0.280	†	0.044	-0.378	0.466	*
p,p'-DDD	0.050	-0.065	0.164		-0.013	-0.074	0.048		0.133	-0.038	0.305		0.186	-0.040	0.413		0.053	-0.059	0.165	†	-0.230	-0.490	0.031	†	
o,p'-DDE	-0.066	-0.190	0.058		-0.008	-0.074	0.058		-0.139	-0.325	0.048		-0.088	-0.336	0.161		0.052	-0.071	0.174		0.142	-0.143	0.428		
p,p'-DDE	-0.032	-0.195	0.131		0.029	-0.057	0.115		†	-0.092	-0.337	0.153		0.017	-0.308	0.343		0.107	-0.054	0.269		0.006	-0.368	0.380	†
o,p'-DDT	-0.097	-0.237	0.042		0.006	-0.068	0.081		-0.215	-0.424	-0.006	*	-0.104	-0.384	0.176		0.110	-0.028	0.248		0.199	-0.123	0.520		
p,p'-DDT	-0.037	-0.213	0.140		0.019	-0.075	0.112		*	-0.064	-0.331	0.202		0.003	-0.350	0.356		0.067	-0.106	0.240	*	0.058	-0.348	0.465	
Dieldrin	-0.189	-0.397	0.019	†	-0.035	-0.145	0.075		*	-0.094	-0.409	0.221		-0.243	-0.661	0.176		-0.149	-0.356	0.057		0.312	-0.168	0.793	
cis-HCE	-0.189	-0.326	0.068		-0.024	-0.127	0.080		*	0.001	-0.297	0.298		-0.092	-0.486	0.302		-0.094	-0.289	0.101		0.139	-0.314	0.592	
HCB	-0.076	-0.373	0.221		-0.112	-0.269	0.046		0.014	-0.433	0.461		0.146	-0.448	0.740		0.132	-0.162	0.426		0.036	-0.647	0.719		
β -HCH	-0.063	-0.246	0.121		-0.009	-0.106	0.089		-0.070	-0.347	0.206		-0.035	-0.402	0.333		0.035	-0.147	0.217		0.101	-0.321	0.524		
Mirex	-0.037	-0.243	0.169		-0.134	-0.243	-0.026	*		-0.207	-0.515	0.100	*	-0.047	-0.456	0.362	*	0.161	-0.042	0.364		0.122	-0.349	0.593	*
Parlar-26	-0.092	-0.235	0.051		-0.039	-0.115	0.037		-0.070	-0.286	0.145		-0.100	-0.386	0.186		-0.031	-0.172	0.111		0.209	-0.119	0.536		
Parlar-50	-0.084	-0																							