



Napieralski, S. A., Buss, H. L., Brantley, S. L., Lee, S., Xu, H., & Roden, E. E. (2019). Microbial chemolithotrophy mediates oxidative weathering of granitic bedrock. *Proceedings of the National Academy of Sciences of the United States of America*, *116*(52), 26394-26401. https://doi.org/10.1073/pnas.1909970117

Peer reviewed version

Link to published version (if available): 10.1073/pnas.1909970117

Link to publication record in Explore Bristol Research PDF-document

This is the author accepted manuscript (AAM). The final published version (version of record) is available online via National Academy of Sciences at https://www.pnas.org/content/116/52/26394.long . Please refer to any applicable terms of use of the publisher.

University of Bristol - Explore Bristol Research General rights

This document is made available in accordance with publisher policies. Please cite only the published version using the reference above. Full terms of use are available: http://www.bristol.ac.uk/pure/user-guides/explore-bristol-research/ebr-terms/

1 Physical Sciences: Earth, Atmospheric, and Planetary Sciences

2 Microbial chemolithotrophy mediates oxidative weathering of granitic

- 3 bedrock
- 4
- Stephanie A. Napieralski^{1*}, Heather L. Buss², Susan L. Brantley³, Seungyeol Lee¹, Huifang Xu¹
 and Eric E. Roden^{1*}
- ¹Department of Geoscience, NASA Astrobiology Institute, University of Wisconsin-Madison,
- 9 Madison WI 53706. ²School of Earth Sciences, University of Bristol, Bristol BS8 1RJ, UK.
- ³Earth and Environmental Systems Institute, and the Department of Geosciences, Pennsylvania
- 11 State University, University Park PA 16802.
- 12
- 13 *Corresponding author e-mail: snapieralski@wisc.edu; eroden@geology.wisc.edu
- 14
- 15 Keywords: Weathering, Chemolithotrophy, Critical Zone

16 The flux of solutes from the chemical weathering of the continental crust supplies a steady supply of essential nutrients necessary for the maintenance of Earth's biosphere. 17 Promotion of weathering by microorganisms is a well-documented phenomenon and is most 18 often attributed to heterotrophic microbial metabolism for the purposes of nutrient acquisition. 19 Here we demonstrate the role of chemolithotrophic ferrous iron [Fe(II)]-oxidizing bacteria in 20 biogeochemical weathering of subsurface Fe(II)-silicate minerals at the Luquillo Critical Zone 21 Observatory in Puerto Rico. Under chemolithotrophic growth conditions, mineral-derived Fe(II) 22 in the Rio Blanco Quartz Diorite served as the primary energy source for microbial growth. An 23 24 enrichment in homologs to gene clusters involved in extracellular electron transfer was associated with dramatically accelerated rates of mineral oxidation and ATP generation relative 25 to sterile diorite suspensions. Transmission electron microscopy and energy dispersive 26 spectroscopy revealed the accumulation of nanoparticulate Fe-oxyhydroxides on mineral 27 surfaces only under biotic conditions. Microbially-oxidized quartz diorite showed greater 28 susceptibility to proton promoted dissolution, which has important implications for weathering 29 reactions in situ. Collectively our results suggest that chemolithotrophic Fe(II)-oxidizing bacteria 30 are likely contributors in the transformation of rock to regolith. 31 32 33 34 35

36

37

38 Significance

We utilized the Luquillo Critical Zone Observatory (LCZO) in Puerto Rico to test the hypothesis 39 40 that mineral-derived Fe(II) within the granitic bedrock at LCZO is capable of supporting microbial Fe(II)-based chemolithotrophy, and that the resultant redox-driven mineralogical 41 42 transformations contribute to bedrock weathering. While this hypothesis had been previously postulated based on theoretical calculations of Fe(II) loss and potential chemolithotrophic Fe(II)-43 oxidizing bacterial growth across the bedrock-saprolite interface, to date it has not been verified 44 45 experimentally. Our study definitively demonstrates the ability of chemolithotrophic Fe(II)oxidizing bacteria to accelerate oxidative transformation of Fe(II)-silicate minerals. In addition, 46 our work presents new insight into the complex microbial community interactions which must be 47 48 considered when considering the role of microorganisms in bedrock weathering.

49

50 **Introduction**

The role of microorganisms in the weathering of minerals has long been recognized (1). 51 More recent interest in the role of Fe(II)-oxidizing bacteria (FeOB) has been driven by the 52 53 recognition that Fe(II)-bearing mineral phases, such as Fe(II)-silicates and pyrite, represent a potential wealth of energy to fuel chemolithotrophic metabolisms, both terrestrially (2) and on 54 55 other rocky planetary bodies such as Mars (3). Thus far the best attempts to characterize the 56 activity of FeOB and their relationship to Fe(II)-silicate weathering come from studies on the subaqueous alteration of the basaltic oceanic crust where it has been demonstrated that FeOB 57 colonize highly reactive basaltic glasses and form thick microbial mats near hydrothermal vent 58 features (4-6). However, controversy remains as to the ability of these marine microorganisms to 59 directly utilize solid phase Fe(II) to fuel their metabolisms (7, 8), and it has been suggested that 60

dissolved Fe(II) released is the major energy source for biomass formation in the vicinity ofhydrothermal vents (9).

63 Compared to the extensive studies targeting oceanic systems, investigations into the role of FeOB in continental weathering are more limited. The potential role of FeOB in Fe-silicate 64 weathering has been postulated, the supposition being that redox driven crystallographic changes 65 66 should be sufficient to lead to mineral dissolution (10). Although it has been established that structural Fe(II) in biotite is capable of supporting FeOB growth in vitro (11), efforts to more 67 fully characterize the role of bacteria in terrestrial weathering processes (10, 12, 13) and to link 68 69 FeOB activity to weathering of volcanic rocks (14) have yielded no definitive evidence for the involvement of FeOB in situ. Nevertheless, multiple lines of circumstantial evidence have been 70 presented for the potential involvement of FeOB in the weathering of the Rio Blanco Quartz 71 Diorite underlying the Rio Icacos watershed of the Luquillo Critical Zone Observatory, Luquillo 72 PR (15-17). The Rio Blanco Quartz Diorite is primarily composed of plagioclase feldspar and 73 74 quartz with lesser amounts of the Fe(II)-bearing silicate phases biotite (~10 wt%) and hornblende (~7 wt%). It is estimated to have one of the highest weathering fluxes known for a granitic 75 material (18). The regolith developed from the Rio Blanco Quartz Diorite consists of a 1-meter 76 77 thick soil overlying an oxidized saprolite zone comprised primarily of quartz, altered biotite, secondary kaolinite, and goethite with a variable depth of 2 m to perhaps 30 m (18, 19). The 78 79 interface of partially altered, fractured rocky material between individual unaltered bedrock 80 corestones and overlying saprolite is termed the "rindlet zone" (20). Here diffusion of oxygen into the crystalline rock is thought to cause oxidation of biotite, producing strain that ultimately 81 82 causes the bedrock to fracture and weather spheroidally, exhibiting a concentric, onion skin-like 83 profile commonly observed during weathering of some granites (21) (see Figure 1 inset). Further

oxidative weathering of biotite occurs within the rindlet zone and the complete depletion of 84 hornblende occurs across a narrow, c.a. 7 cm band of rindlets before the rindlet-saprolite 85 interface (22). Within this zone, an increase in cell density has been previously reported, 86 consistent with theoretical calculations suggesting that the gradient of Fe(II) generated by 87 weathering across this zone is capable of supporting a robust community of lithotrophic FeOB at 88 89 depth (15). Accordingly, we observed an increase in microbial biomass as determined by ATP content of the regolith (Figure 1) relative to the overlying saprolite, indicating the presence of an 90 actively metabolizing microbial community coincident with a sharp gradient in solid-phase 91 92 Fe(II) at the rindlet saprolite interface. Within this biogeochemical framework, we sought to test the hypothesis (15) that mineral-derived Fe(II) is capable of supporting chemolithotrophic 93 cellular growth coupled to Fe(II) oxidation. In addition, electron microscopic analysis and 94 simulated weathering experiments explored how microbial redox driven mineralogical 95 transformations may contribute to the previously documented (15, 18, 20-22) weathering 96 systematics of the Rio Blanco Quartz Diorite. 97

98 **Results and Discussion**

99 Chemolithotrophic Fe(II)-oxidizing enrichment cultures

Ground (<45 μm) Rio Blanco Quartz Diorite was incubated over a period of ca. 2.4 years (864 days) under imposed chemolithotrophic conditions with natural inocula from 3 separate rindlet-saprolite interface samples (Cores A, B and C). Significant oxidation was observed in the presence of a live inocula compared to sterile abiotic controls. The ratio of Fe(II) to Fe(tot) in dilute HCl extracts of solid phase material declined over time from 76.3 to 43.1% (Figure 2a) in the most extreme example of microbial oxidation. This change in the dilute HCl extractable Fe pool corresponds to the oxidation of ca. 0.6% of the total Fe(II) content of the quartz diorite.

107	Given that no significant oxidation occurred under abiotic conditions, our results demonstrate
108	that microbial acceleration of Fe(II)-silicate oxidation was essentially infinite on the time scale
109	of this experiment. ATP abundance, indicating the generation of metabolic energy, was up to an
110	order of magnitude higher in cultures containing diorite compared to cultures provided with pure
111	(Fe(II)-free) quartz sand (Figure 2b,c). Both the quartz sand and the quartz diorite had no
112	detectable (<0.005%) particulate organic carbon (POC) content, suggesting ATP generation was
113	not primarily linked to the oxidation of trace POC in mineral substrates. ATP has been
114	demonstrated to correlate directly with biomass carbon (23, 24). Thus, using a conversion of 10
115	μ mol ATP g ⁻¹ biomass C (25) and assuming that the pool of dilute HCl-extractable Fe(II)
116	represents Fe(II) available for microbial oxidation (11), microbial growth yields in µmol biomass
117	C μ mol ⁻¹ Fe(II) oxidized were estimated. Biomass yields over the first 172 days (to peak ATP
118	production) from individual reactors inoculated with material from Cores A and B were between
119	0.013 and 0.020 μ mol of biomass C μ mol ⁻¹ Fe(II) oxidized (Calculations in SI Appendix Table
120	S1), consistent with previously reported growth yields for neutrophilic chemolithotrophic FeOB
121	in opposing gradient media (26). Growth yields from reactors inoculated with material from Core
122	C were more variable between replicates and higher than would be predicted for Fe(II) oxidation
123	alone. As the extent of oxidation in C reactors was lower than observed for A and B reactors
124	with comparable ATP production, this observation is best explained by input from alternative
125	metabolisms in Core C reactors. However, taken together, these results suggest that the oxidation
126	of mineral-derived Fe(II) in the quartz diorite was the primary source of metabolic energy
127	generation and resultant microbial growth in the majority of reactors. After initial growth,
128	spurred by the availability of fresh mineral surfaces, ATP generation declined across all reactors

129

130

while Fe(II) continues to be oxidized, suggesting the establishment of a maintenance condition, whereby individual cells are still metabolizing without actively increasing in biomass.

- Shotgun metagenomic analysis revealed that the microbial community in the quartz 131 diorite-oxidizing enrichment culture was dramatically simplified compared to the *in-situ* rindlet-132 saprolite sample (SI Appendix Figure S1). The enrichment culture metagenome was dominated 133 134 by organisms belonging to the Betaproteobacteria, including the genera *Cupriavidus* and Burkholderia and the order Neisseriales (Figure SI Appendix S2). Such organisms have been 135 136 previously shown by 16S rRNA gene amplicon sequencing surveys to be abundant in weathering 137 systems (10, 27) and the ability of C. necator to grow by oxidation of Fe-phyllosilicate minerals has been demonstrated (28). Taxonomically, the Neisseriales sp. in enrichments appears to be 138 closely related to the lithotrophic, Fe(II)-oxidizing, nitrate reducing organism 139 Pseudogulbenkiania sp. strain 2002, which is capable of nitrate-dependent growth on solid 140 phase Fe(II) (29). In addition to these dominant organisms having taxonomic affinity to 141 142 previously described FeOB, seven metagenome assembled genomes (MAGs) obtained from the coassembled metagenomes contain homologs to the known Fe(II) oxidation pathway of the 143 acidophilic FeOB Acidithiobacillus ferroxidans (Figure 3). In A. ferroxidans the outer membrane 144 145 bound c-type cytochrome Cyc2 is the iron oxidase (30, 31). As is the case with the oxidation of soluble Fe(II) by A. ferroxidans, the oxidation of mineral bound Fe(II) would necessarily be 146 147 performed extracellularly (11) with subsequent transport of electrons to the intracellular 148 components of the electron transport chain via a periplasmic electron carrier. This process, termed extracellular electron transfer (EET), was originally recognized in dissimilatory Fe(III)-149 150 reducing organisms (32) but has subsequently been shown to be utilized by FeOB (33, 34).
- 151 Homologs to the Cyc2-type EET system have been found to be present in a broad range of FeOB

152 genomes, including those of aerobic neutrophilic FeOB (33, 35) and recently validated via metaomics (36). Organisms of the genera Ralstonia and Rhodopseudomonas which are known to 153 harbor FeOB (37, 38) were among the top 10 genera in the enrichment culture based on read 154 classification, (Figure SI Appendix S2) however, no MAGs containing EET pathways of these 155 genera were obtained. Additionally, while ectomycorrhizal fungi have been noted to oxidatively 156 157 weather structural Fe(II) in biotite (39), fungal associated sequences were not detected in the raw 158 reads for either metagenome or the coassembled metagenome, likely due to the extremely low 159 organic carbon content at the bedrock-saprolite interface (15) and the chemolithotrophic 160 culturing conditions in our experiments.

Many chemolithotrophic organisms are capable of growing autotrophically, most 161 commonly by the use of the ubiquitous enzyme Ribulose-1,5-bisphosphate carboxylase 162 (RuBisCO), which serves as the entry point for inorganic carbon into the Calvin Cycle. Of the 163 MAGs that contained putative EET pathways, three also contained the complete RuBisCo 164 system, including two Cupriavidus MAGs (Figure 3), which supports the idea that these 165 organisms can grow chemolithoautotrophically. Notably, there are multiple MAGs with putative 166 EET pathways that do not contain RuBisCo, including a Xanthomonadaceae, most closely 167 168 related to the soil bacterium Dyella japonica. Though not described as Fe(II) oxidizers, a homolog to Cyc2 gene was also found in the non-autotrophic D. japonica A8 (40). While 169 170 chemolithoheterotrophy is a less common metabolic strategy than chemolithoautotrophy and 171 remains to be validated in Dyella sp., the potential for Fe(II) chemolithoheterotrophy cannot be discounted. Mapping of the metagenomic reads from individual samples back to each MAG from 172 173 the co-assembled metagenome reveals that the putative chemolithotrophs became enriched in the 174 diorite-oxidizing cultures relative to the *in-situ* sample (Figure 3).

175

5 Mineralogical and geochemical effects of FeOB activity

Potential mineralogical changes associated with FeOB activity in the enrichment cultures 176 were assessed via field emission scanning electron microscopy (FE-SEM) and transmission 177 electron microscopy (TEM) with selected area electron diffraction (SAED). Inspection of whole 178 biotite grains revealed a roughening of grain edges after incubation with live inocula that was not 179 180 observed after abiotic incubation over the same time period. Significant alteration of the basal plane of biotite was observed (Figure 4). Etch pits, noted to be formed by siderophore promoted 181 dissolution (41) were not observed on hornblende surfaces. However, microbially oxidized 182 183 hornblende surfaces displayed other subtle differences in morphology suggestive of surface alteration (Figure 5). Upon further inspection via bright field TEM nanosized particles were 184 found along the basal plane of microbially-oxidized biotite and the edges of the surface steps of 185 hornblende (Figure 6). Initial time zero samples from inoculated cultures displayed clean biotite 186 and hornblende crystal surfaces (Figure 6). The lack of these features in the inoculated samples 187 188 at time zero, as well as their absence after 864 days of abiotic incubation (SI Appendix Figures S3,S4), implies that the nano-particles were generated over the course of the experiment by 189 microbial oxidation and not acquired when the weathered inocula were added to the fresh diorite. 190 191 TEM-EDS spectra demonstrate that the nano-particles are Fe-oxyhydroxides, as indicated by Fe enrichment on the microbially oxidized surfaces compared to clean surfaces (Figure 6). The iron 192 oxyhydroxides were around 3-5 nm, similar in size to common examples of ferrihydrite (42), 193 194 suggesting the precipitation of ferrihydrite on the surface of Fe-bearing minerals was triggered by microbial oxidation. 195

In addition to the accumulation of nano-size Fe-oxyhydroxides, small but significant
differences were observed in the total amount of silicon (Si) released from the diorite in the

biotic vs. abiotic reactors. Although the aqueous concentration of Si was indistinguishable
between these treatments (Figure 7), the biotic reactors showed a 13-40% increase (relative to
abiotic controls and time zero samples) in the amount of Si that was released via extraction with
NaOH to raise the pH and desorb any Si that may have been associated with Fe-oxyhydroxides
(43). As a result, there was a significant (two-tailed p=0.0398) increase in total Si release
accompanying the microbial oxidation.

No significant differences in the aqueous concentrations of major cations (Mg, Ca, K, 204 Na) were observed between microbially-oxidized and abiotic or time zero controls. This 205 206 observation is in contrast to numerous studies on microbially-mediated weathering which have demonstrated enhanced release of major rock-forming cations during incubation under 207 heterotrophic conditions (44-46). While initially surprising, it is important to consider the 208 209 mechanistic differences in mineral dissolution under chemolithotrophic versus heterotrophic conditions. It is well known that heterotrophically driven dissolution involves acidolysis and 210 chelation by organic ligands (10). In the absence of respiratory CO₂ generation or low molecular 211 weight organic acids produced as either a byproduct of heterotrophic metabolism or extracellular 212 secretion for nutrient acquisition and/or biofilm formation, one would not expect acidolysis or 213 214 chelation to be the dominant weathering mechanism under the chemolithotrophic, circumneutral pH conditions of our experiments. It has been noted that microscale pH gradients within 215 microbial biofilms on colonized silicate minerals can be lowered as much as 1.1 pH units 216 217 compared to bulk pH (47). Epifluorescence microscopy demonstrated preferential cellular association with solid mineral phases (SI Appendix Figure S5), where cells appeared as sparse, 218 219 singular entities along mineral edges (SI Appendix Figure S6). Similarly diffuse, monolayered 220 biofilms have previously been observed under the carbon limited colonization of basaltic glasses

221 (8). As such, localized biofilm acidolysis is also likely to be insignificant (46). Low molecular weight organic acids generated from the partial oxidation of glucose, in addition to siderophores, 222 also act as effective chelators. Chelation has been noted to be an important driver of silicate 223 dissolution at near neutral pH (48) with several studies noting the effect of siderophores in 224 enhancing solubilization of cations during silicate mineral dissolution (41, 49, 50). Given that 225 siderophores are produced specifically for Fe(III) acquisition as a micronutrient under Fe(III)-226 stress (51), their activity would not be expected to produce the oxidative weathering trend 227 observed in this study. While it is not possible to totally rule out the activity of siderophores in 228 229 this experiment, the data are not consistent with chelation as a primary driver of oxidative weathering under our experimental conditions. Rather, our data collectively point to direct 230 enzymatic oxidation of mineral-derived Fe(II) by chemolithotrophic iron oxidizers for metabolic 231 energy generation. This model is consistent with previously reported models of *in situ* 232 weathering where biological cycling of Fe in the deep saprolite has been inferred based on 233 isotopic measurements (52) and both heterotrophic and lithotrophic microorganisms have been 234 detected at the rindlet-saprolite interface (16, 17). Fe and Mn precipitates previously observed in 235 the outer rindlets, interpreted to result from downward infiltration Fe and Mn rich fluids (53), 236 237 could instead be the result of mobilization and reprecipitation of iron by local oxidative weathering by FeOB in the outer rindlet zone where these organisms are expected to be of 238 importance. 239

240 Enhanced weatherability of microbially-oxidized diorite

Under the imposed chemolithotrophic conditions and considering the proposed
mechanism of a direct enzymatic attack on mineral-derived Fe(II) at circumneutral pH, it follows
that complete stoichiometric dissolution of the Fe(II)-silicate mineral would not immediately

occur and would not be evident over the relatively short time period of this experiment. Rather, 244 Fe(III) may be partially expulsed from the crystal lattices to compensate for the charge 245 imbalance created by oxidation, which would likely result in decreased structural integrity of the 246 mineral as previous studies have shown (11). This mechanism is consistent with the 247 accumulation of nano-sized Fe-oxyhydroxides on biotite and hornblende surfaces (Figure 6). It is 248 249 well noted that crystallographic defects and dislocations are sites of preferential weathering in minerals such as hornblende (54). Thus, it may be envisioned that microbially-oxidized minerals 250 would be more susceptible to other modes of chemical weathering, including proton promoted 251 252 dissolution owing to the inherent disruption of the mineral structure. To address this hypothesis, a portion of the microbially-oxidized quartz diorite was extracted for 24 hours in 10 mM HNO₃, 253 followed by analysis of major cation concentrations in the dilute acid extract measured by ICP-254 OES. HNO₃-extractable Ca and Mg were significantly (two-tailed p=0.0120 and 0.0470, 255 respectively) elevated in microbially-oxidized quartz diorites relative to the unoxidized controls 256 (Figure 7). Major sources of these two cations in the Rio Blanco Quartz diorite include 257 hornblende, biotite and plagioclase (Na,Ca-feldspar). In the case of biotite, which would be the 258 dominant source of K in addition to a source of Mg, HNO₃-extractable K was significantly 259 260 (p=0.0010) lower in biotic reactors than in abiotic controls. It has been shown that the extractability of K from biotite is related to the oxidation state of the octahedral iron with higher 261 K retention correlating to increased oxidation of structural Fe(II) (55-57). It has also been 262 263 observed that oxidized biotites in natural weathering systems can retain significant portions of their K (58). Although K does become depleted (relative to the bedrock) within and above the 264 265 rindlet-saprolite interface (22), this depletion is attributed to continual removal by fluid flow 266 within micro-cracks in the rindlet interiors that form during quartz diorite weathering. Because

267 such fluid flow was absent in our incubation experiments, the repression of K release upon acid extraction observed here is best explained by enhanced retention linked to a decrease in 268 structural Fe(II) in biotite within the closed reaction system. While it is likely that some 269 dissolution of plagioclase contributed to the observed aqueous chemistry, the lack of Fe in its 270 mineral structure makes it generally unresponsive to the activity of FeOB. As such, any 271 272 dissolution of the relatively sodic plagioclase (compared to other rock constituents) upon acid 273 treatment would be expected to be comparable between oxidized and unoxidized diorites. The lack of significant difference (two-tailed p=0.1429) in acid-extractable Na concentrations 274 275 between unoxidized control and microbially oxidized diorites is consistent with this idea and suggests that the difference in acid-extractable Ca, Mg and K between control and oxidized 276 diorites was linked to reduced structural integrity of ferromagnesian minerals as a result of prior 277 FeOB activity. 278

279 **Conclusions**

This study demonstrates that chemolithotrophic FeOB inhabiting the rindlet-saprolite 280 interface of the Rio Blanco Quartz Diorite are capable of growing on mineral-derived Fe(II) as 281 their primary source of metabolic energy, utilizing genomically encoded extracellular electron 282 transfer pathways. The enrichment of these organisms under imposed chemolithotrophic 283 conditions points to their potential to be involved in the subsurface weathering of the Rio Blanco 284 Quartz Diorite. In contrast to the ground quartz diorite used in this experiment, the slow 285 diffusion of oxygen into low porosity fresh bedrock is posited to be necessary for the initial 286 fracturing that forms the rindlet zone (53) and therefore likely modulates weathering over 287 288 geologic time scales. However, once porosity is sufficient to allow advective transport of fluids and microbial colonization along cracks and fractures, in light of the results of this study, it 289

290 seems likely that that FeOB play an important role in the overall weathering regime of the Rio Blanco Quartz Diorite, particularly within the saprolite-adjacent part of the rindlet zone where 291 rapid depletion of mineral-bound Fe(II) is observed. The fact that microbially-oxidized quartz 292 diorites were more susceptible to proton promoted dissolution also has important implications for 293 the effectiveness of acidolysis and/or chelation weathering processes associated with 294 295 heterotrophic microbial metabolism. While the focus of the study was exclusive to the role of FeOB in Fe(II)-silicate weathering and care must be taken when extrapolating laboratory studies 296 to events in natural systems, our findings point clearly to the need for further investigation into 297 298 the interplay between chemolithotrophically and heterotrophically driven silicate mineral weathering. 299

300 Methods

Field Sampling. In June of 2016 three cores (A, B and C) were taken from saprolite atop Guaba 301 Ridge at Luquillo Critical Zone Observatory by hand auger to the depth of refusal (i.e., into the 302 303 outer rindlet zone) which varied from 248 cm (Core B) to 785 cm (Core A), with Core C being of intermediate depth (627 cm) reflecting the topology of the bedrock beneath Guaba Ridge. All 304 cores were taken within close proximity to a previously established lysimeter field, (18, 59) and 305 care was taken to avoid repeat sampling of sites previously cored. Samples were collected 306 aseptically at approximately 40-50 cm intervals for Core A as previously described (15). Cores B 307 and Core C were sampled intermittently. Material collected was shipped overnight to UW-308 Madison on blue ice packs and portions were either refrigerated at 4 °C for live culturing or 309 frozen at -80°C upon arrival for DNA extraction. 0.5 g aliquots of each sample were placed in 20 310 311 mM EDTA and frozen at -80°C for ATP analysis.

312 Chemolithotrophic Enrichment Culturing. Solid phase mineral-oxidizing enrichment cultures were established using whole rock Rio Blanco Quartz Diorite obtained from a road cut exposure. 313 Mineral stoichiometries and abundances were determined previously by White et al. (1998); bulk 314 elemental abundances (aqua regia digestion, ICP-OES analysis, ALS Geochemistry, Reno, NV) 315 are provided in SI Appendix Table 2. Following collection, external weathered surfaces were 316 removed using a rock saw. Large pieces of quartz diorite were fragmented using a jaw crusher to 317 obtain suitable sized fractions for further pulverization using a shatter box. Shattered quartz 318 diorite was sieved to <45 µm. Luquillo artificial groundwater (L-AGW) was prepared to a final 319 320 mM solution concentration of 0.06 MgCl₂·6H₂O, 0.04 KH₂PO₄, 0.05 NaNO₃, 0.1 NaHCO₃, 0.03 Ca(NO₃)₂·4H₂O and 0.01 Na₂SO₄. All glassware was combusted overnight at 550°C to minimize 321 carbon contamination. In an anaerobic chamber, 5.0 g of pulverized quartz diorite or pure quartz 322 sand (Acros Chemicals, 140-381 µm) was placed in a 120 mL bottle and 50 mL anoxic L-AGW 323 was added. Bottles were crimp sealed with a rubber stopper and autoclaved. After sterilization, 324 325 the headspace was flushed with sterile air to render the cultures aerobic. Duplicate reactors of each mineral treatment (quartz diorite or quartz) were inoculated with ca. 1.0 g of material from 326 one of the three (A, B, C) samples obtained from the rindlet-saprolite interface, stoppered and 327 incubated in the dark. Duplicate abiotic controls for each treatment were aerated and left 328 uninoculated. 5.0% (volume) CO₂ was added to the headspace of each bottle as a carbon source 329 330 for autotrophic growth. The pH of reactors after equilibration with CO₂ and mineral phases was circumneutral (6.7-7) in all reactors. Samples were taken immediately following inoculation and 331 after 14, 28, 56, 84, 129, 172, 397 and 864 days. 332

Analytical Techniques. ATP: 0.5 mL of mineral suspension was placed into cold 20 mM

334 ETDA and vortexed and immediately frozen at stored -80°C prior to ATP biomass

335 determination. At the time of analysis samples for ATP were thawed, vortexed once more and centrifuged. ATP content of the supernatant was determined via luminescence using BacTiter-336 GloTM, (Promega, Madison WI) with calibration to a standard curve prepared in 20 mM EDTA. 337 Solid-phase Fe(II): The ratio of Fe(II) to total Fe released by 0.5 M HCl extraction was 338 determined on *in situ* core samples and the solids from 1.0 mL of enrichment culture subsamples. 339 340 The solids were extracted for 24 hours in 5 mL of 0.5 M HCl on an orbital shaker. For natural samples, 0.5 g regolith was added directly to acid for 24-hour extraction. Fe(II) of each extract 341 342 was determined by the standard Ferrozine assay (60) and the measurement was repeated after the 343 addition of hydroxylamine-HCl for determination of Fe(total) with Fe(III) determined by difference. Particulate Organic Carbon: Particulate organic matter of the Rio Blanco Quartz 344 Diorite and Fe(II)-free quartz sand was determined via high temperature combustion using a 345 Flash EA 1112 Flash Combustion Analyzer. Cations: Major cation concentrations (Ca, K, Mg, 346 Na, Si) in the aqueous phase of the cultures were determined using inductively coupled plasma 347 optical emission spectroscopy (ICP-OES) using a Varian Vista-MPX ICP-OES. The aqueous 348 phase from duplicate reactors was pooled, filtered through a 0.22 µm filter and diluted 1:5 in 349 Milli-Q water. Samples were run unacidified to avoid precipitation of silicon, with standards 350 351 prepared for an appropriate calibration curve also in Milli-Q water. Silica: At the termination of the experiment, any sorption of Si to biogenic Fe-(oxy)hydroxides was assessed by high pH 352 desorption. 1.0 mL of culture was aseptically removed and centrifuged to pellet the solids. The 353 354 supernatant was removed and an equal volume of 10 mM NaOH was added to the remaining solids. The slurry was agitated for 24 hours and the supernatant was recovered by centrifugation. 355 356 Si content was determined spectrophotometrically using the heteropoly blue assay. Following 357 verification of consistency between ICP-OES and heteropoly blue Si determination, total Si

release at 864 days was calculated as the sum of aqueous Si and sorbed Si. Epifluorescence
microscopy: Subsamples of live inoculated and abiotic control cultures were taken at 196 days
for epifluorescence microscopy. Whole culture solution was immediately stained with 4',6diamidino-2-phenylindole (DAPI) (ThermoFisher Scientific) following manufacturer's protocols
and imaged on a Nikon E600 compound phase contrast epifluorescence microscope.

363 **Proton Promoted Dissolution Determination.** The susceptibility of oxidized and unoxidized quartz diorites to proton promoted dissolution was assessed as previously described (61) for 364 mineral acid dissolution to avoid ambiguity regarding the potential dual role of organic acids as 365 366 chelators. After 864 days, 1.0 mL of culture from each inocula and the abiotic control were pelleted via centrifugation to recover the solid phase. The supernatant was removed and an equal 367 volume of 10 mM HNO₃ was added and the slurry was agitated for 24 hours on an orbital shaker. 368 The aqueous phase was collected via centrifugation and passed through a $0.22 \,\mu m$ filter. 369 Individual samples where diluted 1:5 in HNO₃ for ICP-OES analysis. Cation concentrations (Ca, 370 K, Mg, Na) were determined by calibration to a standard curve prepared in 10 mM HNO₃. To 371 assess any differences that may have arisen as a consequence of the inclusion of natural 372 weathered material as inocula at time zero, all samples were compared to the initial conditions 373 374 (time zero) for their respective inocula (A, B or C), or fresh diorite in the case of the abiotic control. 375

Mineralogical Analysis. Samples were prepared for field emission scanning electron
microscopy (FE-SEM) by dropping whole, undiluted liquid culture suspensions of time zero, a
microbially oxidized sample inoculated with core A material (the same sample for which the
metagenome was obtained) and an abiotically incubated control onto carbon tape affixed to a
stub mount. Samples were air dried and carbon coated prior to imaging. Images were acquired

381 using a Cameca SXFiveFE with an accelerating voltage of 15 kV. Transmission electron microscopy (TEM) samples were prepared for the same samples as FE-SEM by dropping 382 suspensions of crushed samples onto lacy-carbon-coated 200-mesh Cu grids. TEM imaging and 383 selected-area electron diffraction (SAED) analysis were carried out using a Philips CM200-UT 384 microscope operated at 200 kV in the Materials Science Center at the University of Wisconsin-385 Madison. The chemical composition was obtained using TEM-EDS system equipped with a Li-386 drifted Si detector (Oxford instruments Link ISIS). An electron beam diameter of ~50 nm was 387 used for collecting X-ray EDS spectra. 388

389 **DNA Extraction, Sequencing and Metagenomic Analysis.** DNA was extracted from *in situ* core samples and enrichment culture subsamples via the SDS-based extraction method adapted 390 from Zhou, Bruns, & Tiedje (62). Reagent volumes were appropriately scaled to accommodate 391 392 0.5 g extractions, and 2 volumes of ethanol was used for DNA precipitation at -20° C. Crude DNA was resuspended in 50 µL 10 mM Tris (pH 8). Multiple extractions were performed until a 393 sufficient mass of DNA for metagenomic sequencing was reached. Replicate extracts were 394 395 cleaned and pooled using Zymo Clean and Concentrator-5 (Zymo Research, Irvine CA). Enrichment culture DNA from the 129 day sample was obtained via pelleting 2.0 mL culture and 396 extraction of solids using the same method as above. 397

398 DNA was submitted to University of Wisconsin-Madison Biotechnology Center for 399 metagenomic library preparation and 2x250 paired end sequencing on the Illumina HiSeq 2500 400 Rapid platform. Raw reads were quality trimmed to remove low quality sequences. Taxonomy of 401 individual reads was estimated using Kraken (63) and the standard Kraken database. Reads from 402 individual metagenomic libraries were concatenated and co-assembled using IDBA-UD (64) 403 utilizing the high-performance computing cluster in the Center for High Throughput Computing

(CHTC) at University of Wisconsin-Madison. Assembled contigs were clustered into 404 phylogenetic bins using MetaBAT v2.12.1 (65). The bin set was evaluated for completion and 405 contamination using CheckM (66). Consensus taxonomy of individual bins was determined 406 using single copy housekeeping genes identified in CheckM and MegaBLAST (67) alignment of 407 individual contigs to the Nation Center for Biotechnology nucleotide database using metaWRAP 408 409 (68). Blobology (69) was used to visualize and compare the microbial community compositions. Quantification of the abundance of each bin across samples was performed within the bin 410 quantification module of metaWRAP. Individual bins were reassembled producing a final set of 411 412 metagenome assembled genomes (MAGs) deemed to be of high quality if greater than 70% complete and less than 10% redundant. MAGs were screened for putative extracellular electron 413 pathways as previously described (35). Sequencing data generated in this experiment have been 414 deposited in the Sequence Read Archive (SRA) of the GenBank database under the accession 415 numbers SRR8611926 and SRR8611927, the diorite-oxidizing enrichment culture and in situ 416 417 sample, respectively. Data Analysis. Unpaired t-tests were used in statistical comparison between unoxidized (time 418

- 419 zero and control) and microbially oxidized (A, B, C) using GraphPad Prism version 7.05
- 420 (<u>www.graphpad.com</u>). Two tailed p-values are reported.

421 Data Availability

422 All data discussed in the paper will be made available to readers.

423 Acknowledgements

424 We thank the NSF Luquillo Critical Zone Observatory (LCZO) for access to facilities and

425 assistance with field work. This work was supported by the NASA Astrobiology Institute and a

- 426 University of Wisconsin Microbiome Initiative award to EER. SLB and HLB acknowledge
- support from the LCZO (NSF EAR-0722476 and EAR-1331841).

428 **References**

- J. F. Banfield, W. W. Barker, S. A. Welch, A. Taunton, Biological impact on mineral dissolution:
 application of the lichen model to understanding mineral weathering in the rhizosphere. *Proc Natl Acad Sci U S A* **96**, 3404-3411 (1999).
- W. Bach, K. J. Edwards, Iron and sulfide oxidation within the basaltic ocean crust: implications
 for chemolithoautotrophic microbial biomass production. *Geochim Cosmochim Acta* 67, 38713887 (2003).
- B. M. Jakosky, E. L. Shock, The biological potential of Mars, the early Earth, and Europa. J *Geophys Res Planets* 103, 19359-19364 (1998).
- 437 4. K. J. Edwards *et al.*, Ultra-diffuse hydrothermal venting supports Fe-oxidizing bacteria and 438 massive umber deposition at 5000 m off Hawaii. *ISME J* **5**, 1748-1758 (2011).
- 439 5. C. M. Santelli *et al.*, Abundance and diversity of microbial life in ocean crust. *Nature* 453, 653440 656 (2008).
- 441 6. L. A. Sudek *et al.*, Submarine Basaltic Glass Colonization by the Heterotrophic Fe(II)-Oxidizing
 442 and Siderophore-Producing Deep-Sea Bacterium Pseudomonas stutzeri VS-10: The Potential
 443 Role of Basalt in Enhancing Growth. *Front Microbiol* **8**, 363 (2017).
- 4447.M. Y. Xiong, E. S. Shelobolina, E. E. Roden, Potential for microbial oxidation of ferrous iron in445basaltic glass. Astrobiology 15, 331-340 (2015).
- 8. Bailey, A. Templeton, H. Staudigel, B. M. Tebo, Utilization of Substrate Components during
 Basaltic Glass Colonization by Pseudomonas and Shewanella Isolates. *Geomicrobiol J* 26, 648656 (2009).
- 449 9. A. S. Templeton *et al.*, A seafloor microbial biome hosted within incipient ferromanganese
 450 crusts. *Nature Geoscience* 2, 872-876 (2009).
- 451 10. S. Uroz, C. Calvaruso, M. P. Turpault, P. Frey-Klett, Mineral weathering by bacteria: ecology,
 452 actors and mechanisms. *Trends Microbiol* **17**, 378-387 (2009).
- 453 11. E. Shelobolina *et al.*, Microbial lithotrophic oxidation of structural Fe(II) in biotite. *Appl Environ*454 *Microbiol* **78**, 5746-5752 (2012).
- 455 12. S. Uroz *et al.*, Functional assays and metagenomic analyses reveals differences between the
 456 microbial communities inhabiting the soil horizons of a Norway spruce plantation. *PLoS One* 8,
 457 e55929 (2013).
- 458 13. B. Wild *et al.*, In-situ dissolution rates of silicate minerals and associated bacterial communities
 459 in the critical zone (Strengbach catchment, France). *Geochim Cosmochim Acta* 249, 95-120
 460 (2019).
- 461 14. C. S. Cockell, Life in the lithosphere, kinetics and the prospects for life elsewhere. *Philos Trans A*462 *Math Phys Eng Sci* 369, 516-537 (2011).
- 463 15. H. L. Buss *et al.*, The coupling of biological iron cycling and mineral weathering during saprolite
 464 formation, Luquillo Mountains, Puerto Rico. *Geobiology* **3**, 247-260 (2005).
- 465 16. S. J. Hall *et al.*, Drivers and patterns of iron redox cycling from surface to bedrock in a deep
 466 tropical forest soil: a new conceptual model. *Biogeochemistry* **130**, 177-190 (2016).
- 46717.M. L. Minyard *et al.*, Bacterial Associations with Weathering Minerals at the Regolith-Bedrock468Interface, Luquillo Experimental Forest, Puerto Rico. *Geomicrobiol J* 29, 792-803 (2012).

469 18. A. F. White *et al.*, Chemical weathering in a tropical watershed, Luquillo Mountains, Puerto Rico:
470 I. Long-term versus short-term weathering fluxes. *Geochim Cosmochim Acta* 62, 209-226 (1998).

- 471 19. J. Orlando *et al.*, Architecture of the deep critical zone in the Río Icacos watershed (Luquillo
 472 Critical Zone Observatory, Puerto Rico) inferred from drilling and ground penetrating radar
 473 (GPR). *Earth Surface Processes and Landforms* 41, 1826-1840 (2016).
- 474 20. B. F. Turner, R. F. Stallard, S. L. Brantley, Investigation of in situ weathering of quartz diorite
 475 bedrock in the Rio Icacos basin, Luquillo Experimental Forest, Puerto Rico. *Chemical Geology*476 **202**, 313-341 (2003).
- 477 21. R. Fletcher, H. Buss, S. Brantley, A spheroidal weathering model coupling porewater chemistry
 478 to soil thicknesses during steady-state denudation. *Earth Planet Sci Lett* 244, 444-457 (2006).
- 479 22. H. L. Buss, P. B. Sak, S. M. Webb, S. L. Brantley, Weathering of the Rio Blanco quartz diorite,
 480 Luquillo Mountains, Puerto Rico: Coupling oxidation, dissolution, and fracturing. *Geochim*481 *Cosmochim Acta* 72, 4488-4507 (2008).
- 482 23. D. L. Balkwill *et al.*, Equivalence of microbial biomass measures based on membrane lipid and
 483 cell wall components, adenosine triphosphate, and direct counts in subsurface aquifer
 484 sediments. *Microbial Ecology* **16**, 73-84 (1988).
- 485 24. D. S. Jenkinson, S. A. Davidson, D. S. Powlson, Adenosine triphosphate and microbial biomass in
 486 soil. *Soil Biol Biochem* **11**, 521-527 (1979).
- 487 25. M. Contin, A. Todd, P. C. Brookes, The ATP concentration in the soil microbial biomass. *Soil Biol Biochem* 33, 701-704 (2001).
- 489 26. D. Sobolev, E. Roden, Characterization of a neutrophilic, chemolithoautotrophic Fe(II)-oxidizing
 490 β-Proteobacterium from freshwater wetland sediments. *Geomicrobiol J* 21, 1-10 (2004).
- 49127.C. Lepleux *et al.*, Correlation of the abundance of betaproteobacteria on mineral surfaces with492mineral weathering in forest soils. *Appl Environ Microbiol* **78**, 7114-7119 (2012).
- 493 28. E. Shelobolina *et al.*, Isolation of phyllosilicate-iron redox cycling microorganisms from an illite-494 smectite rich hydromorphic soil. *Front Microbiol* **3**, 134 (2012).
- 49529.K. A. Weber, F. W. Picardal, E. E. Roden, Microbially catalyzed nitrate-dependent oxidation of496biogenic solid-phase Fe(II) compounds. *Environ Sci Technol* **35**, 1644-1650 (2001).
- 497 30. C. Appia-Ayme, N. Guiliani, J. Ratouchniak, V. Bonnefoy, Characterization of an operon encoding
 498 two c-type cytochromes an aa3-type cytochrome oxidase, and rusticyanin in Acidithiobcillus
 499 ferrooxidans ATCC 33020. *Appl Environ Microbiol* 65, 4781-4787 (1999).
- S1. C. Castelle *et al.*, A new iron-oxidizing/O2-reducing supercomplex spanning both inner and outer
 membranes, isolated from the extreme acidophile Acidithiobacillus ferrooxidans. *J Biol Chem* 283, 25803-25811 (2008).
- 50332.D. R. Lovley, D. E. Holmes, K. P. Nevin, "Dissimilatory Fe(III) and Mn(IV) Reduction" in Advances504in Microbial Physiology. (2004), vol. 49, pp. 219-286.
- 50533.R. A. Barco *et al.*, New Insight into Microbial Iron Oxidation as Revealed by the Proteomic Profile506of an Obligate Iron-Oxidizing Chemolithoautotroph. *Appl Environ Microbiol* **81**, 5927-5937507(2015).
- 50834.J. Liu *et al.*, Identification and Characterization of MtoA: A Decaheme c-Type Cytochrome of the509Neutrophilic Fe(II)-Oxidizing Bacterium Sideroxydans lithotrophicus ES-1. Front Microbiol **3**, 37510(2012).
- 511 35. S. He, R. A. Barco, D. Emerson, E. E. Roden, Comparative Genomic Analysis of Neutrophilic
 512 Iron(II) Oxidizer Genomes for Candidate Genes in Extracellular Electron Transfer. *Front Microbiol*513 8, 1584 (2017).
- 51436.S. M. McAllister *et al.*, Validating the Cyc2 neutrophilic Fe oxidation pathway using meta-omics515of Zetaproteobacteria iron mats at marine hydrothermal vents. bioRxiv:10.1101/722066 (2019).

- 51637.E. D. Swanner, R. M. Nell, A. S. Templeton, Ralstonia species mediate Fe-oxidation in
circumneutral, metal-rich subsurface fluids of Henderson mine, CO. Chemical Geology 284, 339-
350 (2011).
- S19 38. Y. Jiao, A. Kappler, L. R. Croal, D. K. Newman, Isolation and characterization of a genetically
 tractable photoautotrophic Fe(II)-oxidizing bacterium, Rhodopseudomonas palustris strain TIE-1.
 Appl Environ Microbiol **71**, 4487-4496 (2005).
- 52239.S. Bonneville, A. W. Bray, L. G. Benning, Structural Fe(II) Oxidation in Biotite by an523Ectomycorrhizal Fungi Drives Mechanical Forcing. Environ Sci Technol 50, 5589-5596 (2016).
- 40. J. W. Chen, K. G. Chan, Genome sequence of Dyella japonica strain A8, a quorum-quenching
 bacterium that degrades N-acylhomoserine lactones, isolated from Malaysian tropical soil. J
 Bacteriol 194, 6331 (2012).
- 527 41. H. L. Buss, A. Lüttge, S. L. Brantley, Etch pit formation on iron silicate surfaces during 528 siderophore-promoted dissolution. *Chem Geol* **240**, 326-342 (2007).
- 529 42. U. Schwertmann, R. M. Taylor, "Iron Oxides" in Minerals in Soil Environments. (Soil Science
 530 Society of America, Madison, WI, 1989), 10.2136/sssabookser1.2ed.c8, pp. 379-437.
- 43. L. Sigg, W. Stumm, The interaction of anions and weak acids with the hydrous goethite (αFeOOH) surface. *Colloids and Surfaces* 2, 101-117 (1981).
- 53344.L. Wu, A. D. Jacobson, M. Hausner, Characterization of elemental release during microbe-534granite interactions at T=28°C. Geochim Cosmochim Acta 72, 1076-1095 (2008).
- 53545.B. Frey *et al.*, Weathering-associated bacteria from the Damma glacier forefield: physiological536capabilities and impact on granite dissolution. *Appl Environ Microbiol* **76**, 4788-4796 (2010).
- W. W. Barker, S. A. Welch, S. Chu, J. F. Banfield, Experimental observations of the effects of
 bacteria on aluminosilicate weathering. *American Mineralogist* 83, 1551-1563 (1998).
- 539 47. L. J. Liermann *et al.*, Microenvironments of pH in biofilms grown on dissolving silicate surfaces.
 540 *Chem Geol* **171**, 1-16 (2000).
- 54148.P. Vandevivere, S. A. Welch, W. J. Ullman, D. L. Kirchman, Enhanced dissolution of silicate542minerals by bacteria at near-neutral pH. *Microb Ecol* 27, 241-251 (1994).
- 54349.B. E. Kalinowski *et al.*, X-ray photoelectron evidence for bacteria-enhanced dissolution of544hornblende. *Geochim Cosmochim Acta* 64, 1331-1343 (2000).
- 545 50. L. J. Liermann, B. E. Kalinowski, S. L. Brantley, J. G. Ferry, Role of bacterial siderophores in 546 dissolution of hornblende. *Geochim Cosmochim Acta* **64**, 587-602 (2000).
- 547 51. J. B. Neilands, Siderophores: structure and function of microbial iron transport compounds. *J* 548 *Biol Chem* **270**, 26723-26726 (1995).
- 549 52. H. L. Buss, R. Mathur, A. F. White, S. L. Brantley, Phosphorus and iron cycling in deep saprolite,
 550 Luquillo Mountains, Puerto Rico. *Chem Geol* 269, 52-61 (2010).
- 55153.A. K. Navarre-Sitchler *et al.*, Porosity and surface area evolution during weathering of two552igneous rocks. *Geochim Cosmochim Acta* **109**, 400-413 (2013).
- 553 54. R. A. Berner, E. L. Sjoberg, M. A. Velbel, M. D. Krom, Dissolution of pyroxenes and amphiboles 554 during weathering. *Science* **207**, 1205-1206 (1980).
- 555 55. I. Barshad, F. M. Kishk, Oxidation of ferrous iron in vermiculite and biotite alters fixation and 556 replaceability of potassium. *Science* **162**, 1401-1402 (1968).
- 55756.R. J. Gilkes, R. C. Young, J. P. Quirk, Artificial weathering of oxidized biotite: I. potassium removal558by sodium chloride and sodium tetraphenylboron solutions. Soil Sci Soc Am J **37**, 25-28 (1973).
- 55957.R. J. Gilkes, R. C. Young, J. P. Quirk, Artificial weathering of oxidized biotite: II. rates of560dissolution in 0.1, 0.01, 0.001M HCl. Soil Sci Soc Am J **37**, 29-33 (1973).
- 561 58. G. Y. Jeong, H. B. Kim, Mineralogy, chemistry, and formation of oxidized biotite in the 562 weathering profile of granitic rocks. *Am Mineral* **88**, 352-364 (2003).

56359.S. F. Murphy *et al.*, Chemical weathering in a tropical watershed, Luquillo Mountains, Puerto564Rico: II. Rate and mechanism of biotite weathering. *Geochim Cosmochim Acta* 62, 227-243565(1998).

- 566 60. L. L. Stookey, Ferrozine-A new spectrophotometric reagent for iron. *Anal Chem* **42**, 778-781 (1970).
- 568 61. C. Balland, A. Poszwa, C. Leyval, C. Mustin, Dissolution rates of phyllosilicates as a function of 569 bacterial metabolic diversity. *Geochim Cosmochim Acta* **74**, 5478-5493 (2010).
- 570 62. J. Zhou, M. A. Bruns, J. M. Tiedje, DNA recovery from soils of diverse composition. *Appl Environ*571 *Microbiol* 62, 316-322 (1996).
- 572 63. D. E. Wood, S. L. Salzberg, Kraken: ultrafast metagenomic sequence classification using exact 573 alignments. *Genome Biology* **15** (2014).
- 574 64. Y. Peng, H. C. Leung, S. M. Yiu, F. Y. Chin, IDBA-UD: a de novo assembler for single-cell and 575 metagenomic sequencing data with highly uneven depth. *Bioinformatics* **28**, 1420-1428 (2012).
- 576 65. D. D. Kang, J. Froula, R. Egan, Z. Wang, MetaBAT, an efficient tool for accurately reconstructing 577 single genomes from complex microbial communities. *PeerJ* **3**, e1165 (2015).
- 578 66. D. H. Parks *et al.*, CheckM: assessing the quality of microbial genomes recovered from isolates, 579 single cells, and metagenomes. *Genome Res* **25**, 1043-1055 (2015).
- 580 67. S. F. Altschup *et al.*, Basic Local Alignment Search Tool. *J Mol Biol* **215**, 403-410 (1990).
- 581 68. G. V. Uritskiy, J. DiRuggiero, J. Taylor, MetaWRAP-a flexible pipeline for genome-resolved 582 metagenomic data analysis. *Microbiome* **6**, 158 (2018).
- 583 69. S. Kumar *et al.*, Blobology: exploring raw genome data for contaminants, symbionts and 584 parasites using taxon-annotated GC-coverage plots. *Front Genet* **4**, 237 (2013).

585 Figure Legends

586 Figure 1 Top: Roadcut exposure of the Rio Blanco Quartz Diorite, used for illustrative purposes

- 587 to conceptualize the subsurface weathering system at Guaba Ridge within the Rio Icacos
- watershed of the Luquillo Critical Zone Observatory in Puerto Rico. The rindlet zone,
- approximately delineated between the solid line (bedrock-rindlet interface) and the dashed line
- 590 (rindlet-saprolite interface), overlies the corestones of bedrock and is the zone of active
- 591 weathering targeted in this study. Scale bar equals 10 cm. The inset at left shows a plan view of
- the rindlet zone exposed elsewhere. Bottom: Total 0.5 M HCl extractable Fe(II) (red circles) and
- 593 ATP content (blue squares) of the actual subsurface regolith obtained by hand auger atop Guaba
- 594 Ridge (Core A) including soil, saprolite and the outer rindlet zone which was partially penetrated
- 595 with auger refusal occurring prior to reaching the bedrock-rindlet interface (note that the
- subsurface rindlet zone is substantially thicker than that revealed by the roadcut). Data points and
- 597 error bars denote the mean and range of triplicate measurements.
- **Figure 2** Molar ratio of Fe(II) to total Fe concentration [Fe(II)/Fe(tot)] in dilute HCl extracts of
- solid phase material in quartz diorite enrichment cultures containing three separate inocula from
- the rindlet-saprolite interface (A,B,C) compared to abiotic uninoculated controls. (a); ATP
- 601 content of cultures containing quartz diorite (b) or quartz sand (c). Data points and error bars
- 602 denote the mean and range of duplicate cultures.
- **Figure 3** Heat map comparison of high-quality MAG abundance (log genomes per million reads)
- and taxonomy between the *in-situ* bedrock-saprolite interface sample (785cm depth) and quartz
- diorite-oxidizing enrichment culture from the same inocula. Stars indicate the presence of

- 606 homologs to the model Cyc2 iron oxidation system of *Acidithiobacillus ferrooxidans*.
- 607 Corresponding gene maps (indicated by star color) are shown for each Cyc2 homolog, compared
- to the model (top). Extracellular or outer membrane putative Cyc2 proteins (green) are scaled to
- the size of the protein with the number of N-terminal heme binding motifs indicated in black and
- 610 C-terminal transmembrane domains in white. Periplasmic electron carriers including monoheme
- c-type cytochromes (blue) or high potential iron-sulfur proteins (dark grey) and hypothetical
- proteins (light grey) are also indicated. Presence of RuBisCo indicated by a circle.
- **Figure 4** FE-SEM images of biotite at the whole grain scale (left) and basal plane (right). Note
- the differences in scale on whole grain images as individual grain sizes are variable. For
- 615 consistency, basal plane images are at the same scale. The approximate area of the basal plane
- presented is outlined in white on the grain scale images. Note the ragged appearance of the basal
- 617 plane observed after microbial incubation.
- **Figure 5** FE-SEM images of hornblende at the whole grain scale (left) and surface scale (right).
- Note the differences in scale on whole grain images as individual grain sizes are variable. For
- 620 consistency, surface images are at the same scale. The approximate area of the hornblende
- 621 surfaces is outlined in white on the grain scale image.
- **Figure 6** Bright field TEM images and SAED patterns (inset) showing widespread nano-sized
- 623 Fe-oxyhydroxide particles (examples indicated by arrows) along the basal plane of microbially-
- 624 oxidized biotite (B) and surface steps of hornblende (F), which were absent in unoxidized time
- ⁶²⁵ zero samples (A and E). The size (ca. 3-5 nm) of the Fe-oxyhydroxides is consistent with
- 626 ferrihydrite. X-ray TEM-EDS spectra confirm the enrichment of Fe (as indicated by the
- 627 Fe/(Si+Al) atomic ratio) on both microbially oxidized biotite (D) and hornblende (H) compared
- 628 to initial time zero surfaces (C and G).
- 629 Figure 7 (a) Concentrations of aqueous, sorbed and total Si released from unoxidized (combined
- time zero and abiotic controls), and microbially-oxidized quartz diorite (A, B and C) after 864
- days of incubation. (b) Concentrations of HNO₃-extractable cations released from unoxidized
- microbially-oxidized samples (A, B and C). For both panels, n=6 for microbially-oxidized
- 633 samples (duplicate cultures from 3 inocula after 864 days), and n=10 for unoxidized (duplicate
- cultures from time zero for 3 inocula and abiotic control and the abiotic control after 864 days
- 635 incubation). Two tailed p-values for unpaired t-test between unoxidized and microbially oxidized
- are indicated for p<0.05 by one star, p<0.01 by two stars, and p<0.001 by three stars.



Fe(II) (µmol g⁻¹)







Microbially oxidized

Abiotic control





Whole grain







Plate Surfaces

Microbially oxidized



LO um

Time zero



Surface













