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Co-constructing future land-use scenarios for the Grenoble region, France

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42 **Highlights**

- 43 • A Participatory Scenario Planning process for downscaling regional normative scenarios.
- 44 • 19 institutions from 6 economic sectors involved throughout a two-year process.
- 45 • Two trend and two break-away scenarios with storylines and projected land cover.
- 46 • Three spatial models to project land use change by 2040 at 15 m resolution.
- 47 • Multi-scale participatory normative scenarios for supporting land planning.

48 **Abstract**

49 Physically and socially heterogeneous mountain landscapes support high biodiversity and
50 multiple ecosystem services. But rapid landscape transformation from fast urbanisation and
51 agricultural intensification around cities to abandonment and depopulation in higher and more
52 remote districts, raises urgent environmental and planning issues. For anticipating their future
53 in a highly uncertain socio-economic context, we engaged stakeholders of a dynamic urban
54 region of the French Alps in an exemplary interactive Participatory Scenario Planning (PSP)
55 for co-creating salient, credible and legitimate scenarios. Stakeholders helped researchers
56 adapt, downscale and spatialize four normative visions from the regional government, co-
57 producing four storylines of trend versus break-away futures. Stakeholder input, combined
58 with planning documents and analyses of recent dynamics, enabled parameterisation of high-
59 resolution models of urban expansion, agriculture and forest dynamics. With similar
60 storylines in spite of stakeholders insisting on different governance arrangements, both trend
61 scenarios met current local and European planning objectives of containing urban expansion
62 and limiting loss and fragmentation of agricultural land. Both break-away scenarios induced
63 considerable conversion from agriculture to forest, but with highly distinctive patterns. Under
64 a commonly investigated, deregulated liberal economic context, encroachment was random
65 and patchy across valleys and mountains. A novel reinforced nature protection scenario
66 affecting primarily mountain and hilly areas fostered deliberate consolidation of forested areas
67 and connectivity. This transdisciplinary approach demonstrated the potential of combining
68 downscaled normative scenarios with local, spatially-precise dynamics informed by
69 stakeholders for local appropriation of top-down visions, and for supporting land planning and
70 subsequent assessment of ecosystem service trade-offs.

71 **Key words**

72 Participatory scenario planning, Scenario downscaling, Land use and land cover modelling,
73 Landscape conversion, Mountain regions

74 **1. Introduction**

75 Societies are realising the ecological limits to socio-economic development (Griggs et al.
76 2013; Steffen et al. 2015). There is at the same time increased recognition of the benefits that
77 ecosystems can provide for society (Díaz et al. 2015). Nature's benefits and 'Nature-Based
78 Solutions' are seen as supporting future socio-economic development, including in developed
79 countries (Maes and Sanders 2017; Nesshöver et al. 2017), requiring changes in social values
80 and governance (Colloff et al. 2017; Kabisch et al. 2016). Consistent with this movement,
81 ecological insights, and specifically ecosystem services assessments are increasingly
82 incorporated into land use planning (Albert et al. 2014; Cabral et al. 2016; Opdam et al. 2015
83 Turkelboom et al., 2017). This poses challenges to planners and decision makers for bringing
84 ecosystem services into political agendas, building their knowledge and capacity, and
85 producing relevant, salient and legitimate assessments of the sustainability of land plans
86 (Albert et al. 2014). Participatory scenario planning is one of the tools to achieve this
87 (Rounsevell et al. 2012).

88 Scenarios, defined as coherent and internally consistent descriptions of the future (Alcamo
89 2009), allow exploring a range of plausible futures without gaging their probability (Peterson
90 et al. 2003). By exploring multiple alternative futures and exploring key uncertainties on
91 drivers and their impacts (Kok et al. 2007; Peterson et al. 2003; Rosa et al. 2017), exploratory
92 scenario planning promotes understanding of complex systems dynamics (Carpenter et al.
93 2009), and expands thinking horizons of scientists, stakeholders and decision makers. As such
94 scenario processes foster creative solutions to environmental problems (Biggs et al. 2007;
95 Peterson et al. 2003). In planning, normative approaches focusing on desired futures may be
96 preferred to exploratory approaches because of their greater saliency and legitimacy (Albert et
97 al. 2014, Castella et al. 2014). Normative, or target-seeking scenarios (Rosa et al. 2017)
98 complement exploratory scenarios by exploring desired scenarios and comparing them to
99 undesired ones to support the design of pathways towards preferred futures (Lavorel et al.,
100 2019; Hanspach et al., 2014; Nieto-Romero et al., 2016; Oteros-Rozas et al., 2013; Palomo et
101 al., 2011). Their value has recently been emphasised for empowering stakeholders in global
102 change adaptation and for fostering institutional and social learning (Sharpe et al. 2016, van
103 Kerkhoff et al. 2018, Lavorel et al. 2019).

104 Among scenario methods, participatory scenario planning (PSP) is defined as engaging
105 stakeholders along with scientists at various stages of the scenario development process
106 (Oteros-Rozas et al. 2015). PSP is increasingly used in environmental research including for
107 analysing global change impacts (Harrison et al. 2015; Moss et al. 2010) or sustainable
108 futures (Bohunovsky et al. 2011; Nieto-Romero et al. 2016). PSP has been used in ecosystem
109 service (ES) research to integrate quantitative, and sometimes spatially-explicit ES
110 assessments with stakeholder demand (see overviews and examples in Albert et al. 2014;
111 Oteros-Rozas et al. 2015; Plieninger et al. 2014). Beyond usual benefits of scenario planning,
112 PSP combines multiple sources of academic, political and civil knowledge, and fosters
113 dialogue and social learning. In the case of environmental issues, PSP aims to foster
114 communication, planning and cultural change from sectoral to trans-sectoral policy, planning
115 and management.

116 In the last decade, land use and ecosystem service PSP has gained currency from local
117 (Hanspach et al. 2014; Oteros-Rozas et al. 2013; Palomo et al. 2011; Plieninger et al. 2013;
118 Schirpke et al. 2017), to national or regional (Cradock-Henry et al., 2018; Mitchell et al. 2015;
119 Reed et al. 2013) and to continental scale (Harrison et al. 2015, Verkerk et al., 2018).
120 However, in spite of its critical role for policy and decision-making, ecosystem service PSP
121 has been significantly less used at sub-national regional than at landscape or municipality
122 scales. Further, multiscale scenarios add to single-scale scenarios by combining a top-down,
123 expert-led component to identify and downscale larger scale scenarios, and a bottom-up
124 participatory process that provides local expertise on specific conditions, especially social,
125 and spatial aspects (Kok et al. 2017). Developing practice in participatory multiscale
126 scenarios (Kok et al. 2007; Lamarque et al. 2013) opens avenues for producing salient and
127 relevant scenarios for regional land use planning.

128 The Grenoble region, in the French Alps, is a typical European urban region facing issues of
129 development in a context of high environmental and amenity values and with high spatial
130 diversity (Vannier et al. 2016). The region's agriculture depends on future policy and social
131 orientations, also on climate and ecosystem changes and adaptations. Future local and
132 external demands for recreation and tourism add to uncertainties to be incorporated into future
133 scenarios (Brunner et al. 2017; Kohler et al. 2017). A broad institutional and citizen
134 participatory urban planning process took place from 2008 to 2012 to produce a development
135 plan (SCoT – Schéma de Cohérence Territoriale) towards 2030, aiming to reconcile a
136 spatially balanced economic growth and environmental objectives, especially from recent
137 French climate and biodiversity legislation and policy. In this context, the objective of this
138 study was to showcase a highly participatory scenario downscaling approach for developing
139 with local decision-makers high-resolution spatially-explicit land-use scenarios. The final
140 outcome is a subsequent assessment of planning alternatives for future ecosystem services
141 trade-offs. We aimed to develop an exemplary participatory scenario process relevant to
142 similar urban regions in developed mountain and other regions, meeting the following criteria:
143 (i) relevance to the specific issues of the study area, as outlined by the current land plan and as
144 expressed by stakeholders; (ii) consistent with larger scale socio-economic scenarios through
145 downscaling; (iii) spatially-explicit.

146 This paper presents four steps for co-producing downscaled normative scenarios using a
147 combination of qualitative and quantitative methods with extensive stakeholder participation:
148 1) scoping of pre-existing visions and scenarios, 2) refining and spatializing scenarios with
149 stakeholder to produce storylines, 3) projecting and 4) analysing land use change at the
150 regional and municipality scale, and consequences for landscape patterns. We argue for the
151 generic advantages of this participatory downscaling methodology and end with considering
152 scenarios implications for future land planning and ecosystem services provision.

153 2. Methods

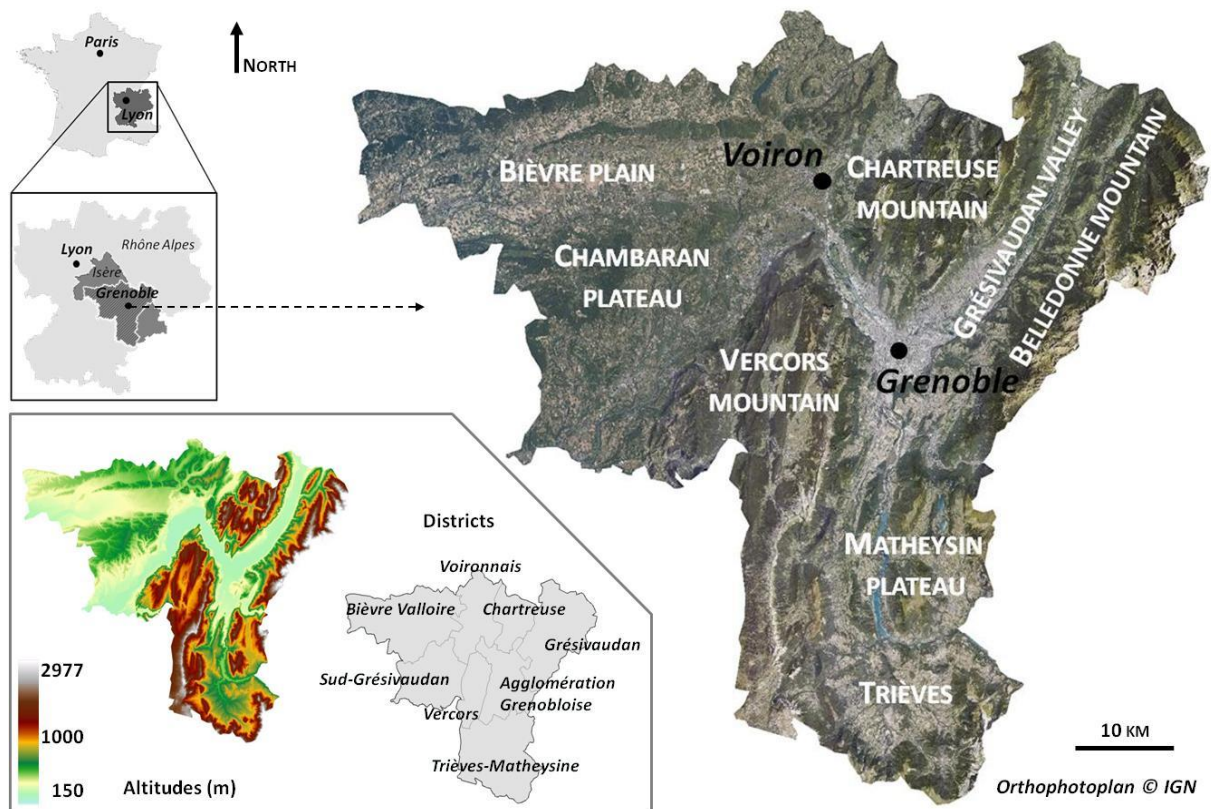
154 2.1. Study site

155 Grenoble is one of the most active and dynamic French metropolitan areas. With an extent of
156 4450 km², the Grenoble urban area hosted in 2012 around 800,000 inhabitants. Our study
157 encompasses the area of economic influence of Grenoble, especially regarding employment.
158 With highly diverse physical and natural characteristics, all significant landscape units in an
159 Alpine region, plains, plateaus and mountains are represented, resulting in contrasted and
160 heterogeneous landscapes (Figure 1). The region is structured by three mountain ranges:
161 Vercors, Chartreuse and Belledonne, culminating at 2977m. River valleys favour urban
162 sprawl, as well as to a lesser extent the Bièvre plain. Mountain ranges benefit from a wide
163 range of protection measures with two natural parks and several conservation areas. Most of
164 the 311 municipalities within 50 km of the city of Grenoble are integrated into the Grenoble
165 SCoT¹ planning area (Schéma de Cohérence Territoriale), whose primary aim is to contain
166 urban expansion and preserve natural assets while supporting equitable economic
167 development at the scale of a small region. For a spatially-explicit specification of scenarios,
168 we considered eight districts regrouping municipalities according to their biophysical features
169 and broad land planning districts (Figure 1).

170 Recent land use trends are consistent with other European mountain regions. Between 1998
171 and 2009 urban use spread at the expense of agricultural land (29 km², + 7% over the 11 year
172 period), either in the valleys near Grenoble or in agricultural plains (Vannier et al. 2016). This
173 expansion was nearly exclusively a densification of urban patches or adjacent to existing
174 urban areas, complying with current urban planning. Agricultural land-use remained stable,
175 mostly because it is largely determined by the physical geography of the study site with
176 permanent grasslands dominant above 800-1000 m altitude, while broad acre crops are
177 preferentially located in the valley bottoms and plains; landscapes on plateaus and hilly areas
178 comprise mosaics of grassland and spring crop successions (Lasseur et al., 2018). Other forest
179 and semi-natural areas also remained stable.

180

1 The SCoT, Territorial Coherence Scheme is a French planning document that determines, for groups of municipalities, common objectives for urban planning, housing, transport, and business and retail areas.
<http://www.region-grenoble.org/index.php>



181
 182 Figure 1 – Grenoble urban area: location map, districts and physical geography.

183

184 **2.2. Participatory scenario process**

185 In order to produce quantitative and spatially-explicit scenarios fitted to regional challenges
 186 and incorporating social, economic and governance dynamics, we developed a
 187 transdisciplinary process involving an interdisciplinary team of eight researchers along with
 188 nineteen stakeholders from the main decision and land management sectors over two years
 189 (2014-15). Researchers in biodiversity, urban planning, agronomy and forestry were involved
 190 through individual consultations and workshops. The nineteen stakeholders were involved in
 191 land management of the Grenoble area within the local government, management
 192 organisations or NGOs and represented, albeit not exhaustively, the predominant land
 193 planning, agriculture, forest, tourism, nature conservation and water management sectors
 194 (Supplementary table 1). They were part of the advisory committee established at the
 195 beginning of the research process (spring 2013) and selected among collaboration networks of
 196 researchers and through snow balling based especially on recommendations from the land
 197 planning agency and the local government (Bierry & Lavorel 2016). We note that the private
 198 and industrial sectors were not represented due to unsuccessful contacts during project
 199 initiation, and likely to their self-perceived less direct role in land management and planning.

200 Through three steps combining qualitative and quantitative methods (Figure 2), this process
 201 aimed to describe alternative visions by 2040 and to translate their socio-economic and
 202 governance characteristics into land use projections. Although the development plan targeted
 203 2030, the 2040 horizon was chosen first for consistency with the strategic horizon at the larger

204 regional (NUTS2, Rhône-Alpes) scale and second to consider more pronounced climate
205 change impacts.

206 As a first step (January 2014) researchers scoped pre-existing local, national and international
207 land-use and/or biodiversity visions and scenarios (Supplementary table 2) and their strengths
208 and weaknesses regarding the project's objectives. The Montagne 2040 visions (Centre
209 Economique, Social et Environnemental Régional Rhône-Alpes, 2013) were selected as most
210 relevant and legitimate, especially given their focus on mountain challenges, which were not
211 considered in larger-scale scenarios, and their familiarity to many stakeholders. These visions
212 were the outcome of a complex two-year expert process led by the Rhône-Alpes
213 administration region interrogating its development pathways given climate change, regional
214 natural and human capital and the vulnerability of mountain economies. We analysed their
215 context scenarios and four final storylines, and identified key driving variables such as the
216 availability and access to natural resources. Through this process we translated the storylines
217 as visions for the Grenoble region considering its biophysical and socio-economic
218 specificities. These four scenarios documented main socio-economic orientations and their
219 local translation in terms of governance, socio-economic dynamics and key activities
220 (agriculture, forestry, water, recreation and tourism, nature conservation and land planning),
221 land use and expected impacts on natural resources, and were summarised as a poster.

222 The second step aimed to produce refined qualitative, locally-specific and spatially explicit
223 translations of the four scenarios by incorporating actors' knowledge of local issues and of
224 social and ecological dynamics. We first aimed to critique the realism of the Montagne 2040
225 visions, originally designed for thought-provoking contrasts and not aimed for impact
226 projections, and their local applicability. Second we aimed to adapt their region-wide socio-
227 economic settings and institutions to the local context. Third we aimed to downscale the
228 visions for the eight districts and for different socio-economic activities using qualitative,
229 semi-quantitative and spatial information.

230 A one-day workshop (March 2014) attended by the nineteen stakeholders was facilitated by
231 four researchers and a professional facilitator. Stakeholders were responsible for managing
232 discussions within each session and for presenting collective conclusions. After an
233 introductory presentation of objectives and of the Montagne 2040 approach, already familiar
234 to many participants, participants were allocated to four groups, each with representation of
235 socio-economic sectors. A researcher presented one scenario per group and coordinated a
236 discussion on its local relevance, its main directions and limits. This discussion was supported
237 by the poster from step 1. During the first session groups were tasked with describing
238 ecosystem services demand for their scenario. The second session brainstormed the associated
239 governance. Following these two sessions, each group presented in plenary their respective
240 scenario and discussion outcomes so as to familiarise all participants with all four scenarios
241 and their local adaptation. During the third session stakeholders were allocated to four
242 geographic groups (each comprising two similar, adjacent districts) and successively analysed
243 the four scenarios to specify land use and management (from the basis of the Land Use and
244 Land Cover -LULC- map described in section 2.3.1. and Supplementary table 3), and their
245 allocation across the eight districts using drawings and/or notes on maps. After a presentation

246 of each group's results a final plenary discussion addressed the relevance of the resulting
247 scenarios. This resulted in a final collective choice of directions for the project's scenarios.
248 The transcription and the analysis of the workshop's results produced four locally-adapted
249 and downscaled scenarios including a description of the socio-economic context, a qualitative
250 specification of land use and management, along with semi-quantitative and spatially-explicit
251 information.

252 The third step aimed to quantify the scenarios in a spatially-explicit fashion. We combined the
253 workshop storylines and maps with a detailed analysis of planning and policy documents and
254 of public and research reports (Supplementary table 2). The SCoT, which quantifies and
255 specifies location of planning objectives, was the main document used as a starting point to
256 translate scenarios into quantity and location, complemented by the management plans of the
257 Vercors and Chartreuse regional parks. Their specifications were applied directly for the
258 Business as usual scenario and were adjusted for the other three scenarios according to
259 stakeholder input during the workshop. To quantify these adjustments which were often at
260 best semi-quantitative, researchers combined workshop and SCoT data with an analysis of
261 land-use trends since 1998, expected climate impacts (following Intergovernmental Panel on
262 Climate Change scenario RCP 8.5), local interdisciplinary scientific expertise (ecology,
263 agronomy, forestry, economics) and ad hoc in depth interviews with key stakeholders (e.g.
264 land planners, regional government) to determine quantitative land allocation rules. Detailed
265 storylines describing the socio-economic and governance context, its translation into
266 economic activities and land-use projections were the output.

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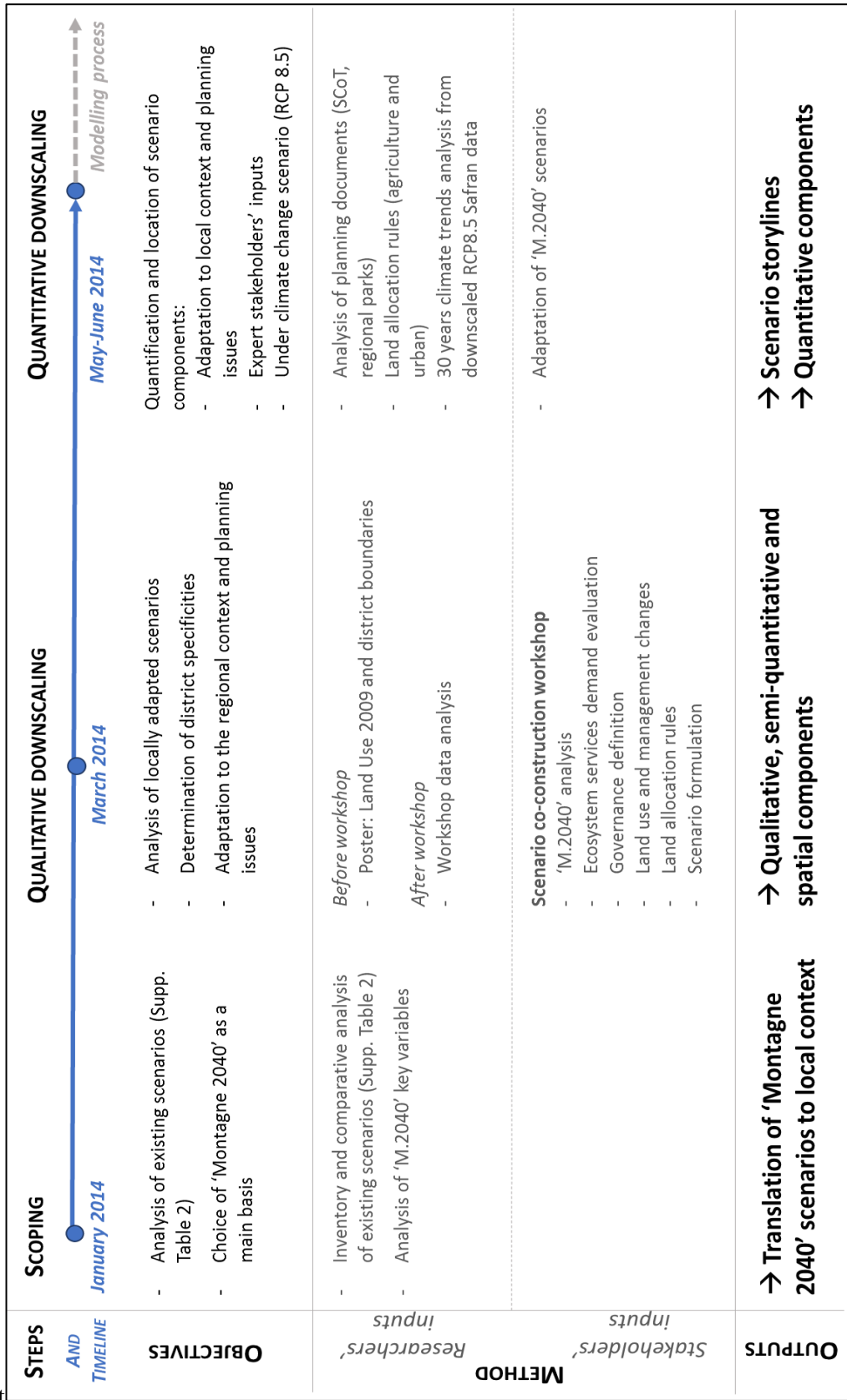


Figure 2 - Three steps for the participatory production of four locally relevant, spatially-explicit scenarios.

271 **2.3. Scenario modelling**

272 **2.3.1. Analysis of current and past landscape dynamics**

273 To model future land use under the four scenarios we analysed main changes in terms of
274 amounts and spatial allocation over the 1998-2009 period. A detailed description of the data
275 sets and analyses are provided by Vannier et al. (2016). Briefly, maps at a 1/15000 scale for
276 23 Land Use and Land Cover (LULC) types nested at three levels were produced for 1998,
277 2003 and 2009 using a multi-source approach (Supplementary table 3). These were refined for
278 agricultural land by characterising 5-year crop type / grassland successions at parcel scale
279 (Lasseur et al., 2018). The resulting maps, with 41 LULC classes at two levels
280 (<http://www.projet-esnet.org/en/cartes/>), were analysed with a particular emphasis on urban
281 spread dynamics, agricultural geographical patterns and land abandonment / forest regrowth.

282 **2.3.2. Land use modelling**

283 Our modelling framework operated at two spatial scales, the entire site and its eight districts
284 (Figure 1). Simulations were run at a 5-year time step for a total of 30 years. We incorporated
285 governance levels from the EU (e.g. the Common Agricultural Policy determining viability of
286 mountain agriculture), to national (e.g. nature protection legislation determining zoning of
287 protected areas), and regional or local (e.g. land planning constraining urban development).
288 Land-use projections were modelled at the finest available scale, e.g. the parcel for
289 agriculture, and forest or urban patches. As the analysis of recent landscape changes revealed
290 three major types of landscape dynamics for urban, agricultural, and forested and semi-natural
291 areas respectively, we developed three distinct models for urban spread, agricultural land and
292 forest expansion. To achieve this, numerous types of spatial, statistical, existing data were
293 used (Supplementary table 4) for model parametrisation (Supplementary table 5).

294 *Urban spread*

295 Urban spread is the most rapid process in the study area. Over periods of five years numerous
296 but rather small patches are converted. The overall transfer from (mostly) agricultural land to
297 urban areas is rather small, but this large number of new patches requires careful modelling in
298 order to obtain realistic results. Two different types of processes were distinguished: the
299 creation of new residential areas, and the creation of new industrial and commercial areas.

300 We used the spatially-explicit statistical modelling platform Dinamica EGO to construct our
301 urban spread model (Soares-Filho *et al.* 2013). Transition probabilities were obtained from
302 historical data through the statistical correlations of past changes (from Vannier et al. 2016)
303 with spatially-explicit predictors. From an initial list of 18 such parameters, including
304 geographical (e.g., slope) and socio-economic data (e.g., cost of real estate, employment rate
305 at the municipality level), we retained four geographical parameters sufficient to capture most
306 of the historical urbanization trends: altitude, slope, distance to existing urban areas, distance
307 to roads. The statistical relevance of potential predictors was assessed through Cramer tests;
308 their statistical independence through Cramer tests, correlation and principal component
309 analysis. Finally, the overall quantity and location of LULC transitions were specified for the
310 whole study area per time step, using calibration from historical data and trends specified by
311 stakeholders and land planning documents.

312 *Agricultural land*

313 Types of dynamics were established regarding the scenarios and quantification of dynamics
314 was estimated regarding the past dynamics on each district. Changes in the area of agricultural
315 land result from two distinct mechanisms. First boundary changes reflect the loss of
316 agricultural land due to urban extension, or agricultural abandonment leading to forest
317 regrowth. The former was simulated through the urban spread model. The latter varied across
318 scenarios in terms of amounts and location. The historical analyses of limited change revealed
319 a preferential abandonment of small parcels adjacent to forest and sloping and depending on
320 altitude (Vannier et al., 2016), whereas in scenarios of massive abandonment we targeted
321 either specific crop succession types or areas adjacent to forests of green corridors.
322 Abandoned parcels were allocated to the “transition” land-cover type (Supplementary table
323 3).

324 Second changes in agricultural practices leading to changes in crop succession within the
325 agricultural area were addressed with a spatial GIS model. The agricultural practices were
326 drawn from a database of crop successions and an analysis of agricultural statistics
327 respectively (Supplementary table 3 and 4). Scenario defined which crop successions were
328 targeted for change, the amount of change per succession type, per district, and spatial
329 allocation rules. For instance in the Business as usual scenario, in the Vercors district 3% of
330 current grassland-dominated successions will incorporate a crop by 2040. Fields were targeted
331 for change in agricultural succession depending on spatial allocation rules (proximity,
332 distance, random effects etc.) drawn from the storylines and additional documents. Type of
333 changes in agricultural succession were also influenced by projections of climate impacts
334 (Ruget et al. 2013).

335 *Agricultural abandonment and forest regrowth*

336 The model of woody encroachment and forest regrowth starts from the projections of
337 abandoned parcels by the agricultural land model (allocated to “transition” class,
338 Supplementary table 3 and 5). Transition to forest regrowth depends on altitude (<800 m,
339 800-1200 m, 1200-1500 m, >1500 m), district, nearby forest type (broadleaf, conifer, mixed
340 forest or shrubby heathland) and time since abandonment (10-20 year-old forest, 25-30 year-
341 old forest, 20-30 year-old woody heathland at higher altitude). The analysis of dynamics
342 between 1998 and 2009, additional data concerning forest regrowth from 1993-1997 and
343 farmer interviews in 2012-2014 (Supplementary table 4) allowed us to identify areas prone to
344 woody encroachment and the temporal dynamics of forest recolonization. The type of
345 recolonizing forest was determined from analyses of BD Topo data and of sylvo-ecoregions
346 (Supplementary table 4). Climate change impacts were considered to already be current,
347 whereas more drastic impacts on forest dynamics and management would not be expected
348 until the second half of the 21th century (e.g. Seidl et al. 2011).

349 The agricultural and forest models were implemented using ArcGis model builder (version
350 10.2, ESRI Inc.).

351 **2.4. Analysis of model outputs - indicators**

352 The 2009 LULC map and its projections for 2040 were analysed in three steps. First, site-
353 level percentages of LULC changes documenting overall dynamics of the six main classes
354 under the scenarios. Second, municipality-level indicators summarizing relevant information
355 for managers and decision-makers were computed. We aggregated the six main land cover
356 classes (Supplementary table 3) to municipality and district scale for 2009 and 2040
357 projections and analysed their changes graphically. Third, landscape metrics were computed
358 at the finest available map resolution documenting changes in overall spatial structure with
359 relevance to spatially sensitive ecosystem services (Verhagen et al. 2016). We quantified
360 landscape heterogeneity, texture, and graininess based on area, patch number (NP), mean
361 patch size (MPS) at LULC class level; and Shannon Diversity Index (SHDI) at landscape
362 level (Cushman et al., 2008), using Fragstats® (McGarigal et al., 2012) for the 1998-2009
363 (observed) and 2009-2040 (projected) periods. This landscape metrics analysis focused on the
364 three classes undergoing most of the changes: urban, agricultural and forested areas.

365 **3. Results**

366 **3.1. Storylines and scenario parameterisation**

367 Four descriptive and quantitative plausible scenarios were produced, with two scenarios based
368 on current trends and two break-away scenarios.

369 **Business as usual (BAU):** A local implementation of the Montagne 2040 Business as usual
370 scenario. Based on currently existing policy and planning documents (the SCoT and regional
371 natural park (PNR) management plans), development in this scenario is based on current
372 regional planning and management policies. Learnings from an analysis of past dynamics are
373 taken into account so as to maintain coherence with current trends and take into account the
374 coordinated policy objectives of the Grenoble urban region.

375 **Local development:** A variant of the Business as usual scenario not considered in Montagne
376 2040, and not captured by its local green development vision. While like Business as usual
377 this scenario is based on the continuation of current dynamics, the objectives of decentralised
378 development at the regional level such as prescribed in the SCoT are not adopted. Instead,
379 new governance arrangements with greater local control on land allocation and strengthened
380 authority for protected areas favour focused development around selected urban centres,
381 reinforcing their attractiveness and densifying contiguous urban expansion. In line with
382 current policies for sustainable development and the preservation of natural areas, the
383 emphasis is placed on local regional development via economic activity and tourism,
384 favouring local marketing (timber, agriculture) and reinforcing regional natural parks.

385 **Rewilding:** A local implementation of the Montagne 2040 corresponding vision. This
386 scenario replaces current policies with a strong nature conservation orientation, placing
387 natural areas and in particular mountain areas in strict reserves. Consequently population and
388 economic activities decrease drastically in these areas and are transferred to lowlands. The
389 handicaps linked to the lack of use of these areas, and the overall reduced economic

390 attractiveness of the region exacerbate their gradual abandonment and promote forest
391 encroachment, while increased urbanisation and the development of currently existing
392 economic activities are concentrated in valleys.

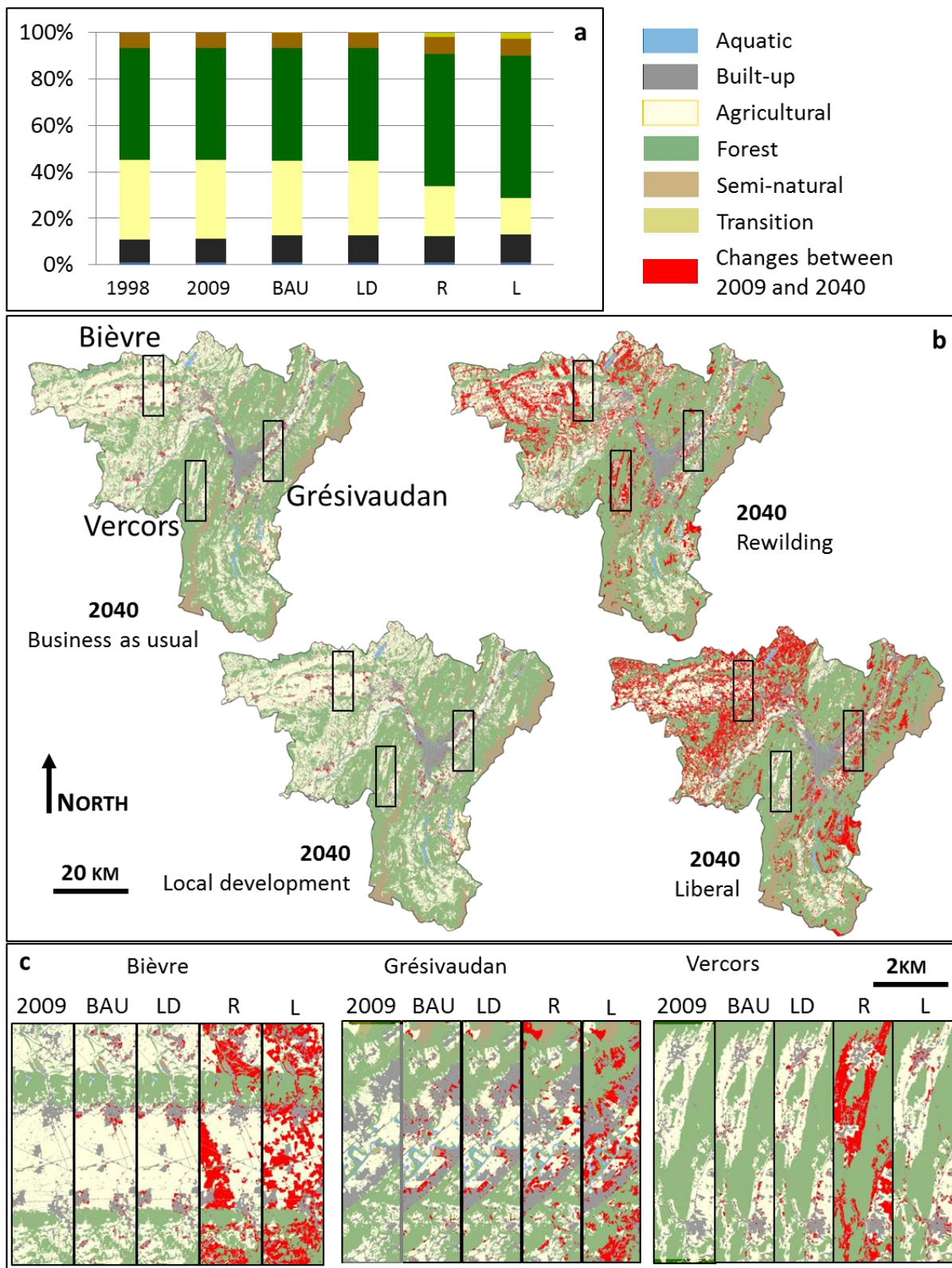
393 **Liberal:** An adapted implementation of the Montagne 2040 ultra-liberal vision which focused
394 strongly on tourism. This scenario breaks away from current policy with a marked
395 liberalisation of public policies, development driven by private investment, and thus major
396 social and economic divides. The urban / rural divide is reinforced, accentuating disparities in
397 access to resources, housing and services, as well as inequities regarding management of
398 natural hazards. Market liberalisation and the absence of land-use regulation via public
399 policies is detrimental to local agriculture: agricultural landscapes and practices undergo
400 major modifications, and their area is reduced by urban expansion. Mountain areas are also
401 affected, with development tied to attractiveness for tourism activities.

402 **3.2. Overall land-use / land-cover changes**

403 LULC maps for each scenario are presented in supplementary figure 1. Given minimal overall
404 change under the BAU and Local development scenarios (Figure 3a), corresponding maps
405 were quite similar to 2009. Urban spread ($> +10\%$) around current urban areas was the major
406 change under the BAU and Local scenarios, primarily at the expense of agricultural land
407 (Figure 3a). In contrast, the two break-away (Rewilding and Liberal) scenarios showed large
408 overall change, with contrasting spatial patterns (Figure 3a, b). They resulted in considerable
409 forest expansion ($+20\%$ and $+30\%$ respectively) at the expense of agricultural land (-35%
410 and -52% respectively). Under Rewilding forest expanded along already existing corridors,
411 thus reinforcing initial spatial patterns. In the Liberal scenario land abandonment was
412 randomly distributed in less productive areas. Given considerable encroachment, forests then
413 become spatially continuous in less productive areas. For all four scenarios 90% of the
414 changes concentrated below 1000m altitude (Figure 3b). Changes thus affected most strongly
415 the Bièvre plain and the Grésivaudan valley – especially under trend scenarios, as well as
416 hilly areas around Voironnais, and the Chambarans and Matheysine plateaus under the break-
417 away scenarios. Conversely lower areas in Trièves appeared stable under all scenarios. The
418 Rewilding scenario specifically affected mountain areas (20% of the total changes).

419 A detailed examination of the most dynamic areas (Bièvre, Grésivaudan and Vercors, Figure
420 3c) highlights minor changes between 2009 and both trend scenarios. Strong urban spread
421 concentrated in the plains and valley bottoms with relatively less impacts in Vercors. The
422 increased density of green corridors in plains constitutes the major landscape change by 2040
423 in the Local scenario along the edge of the Bièvre intensive agricultural area, and along the
424 bottom of the Grésivaudan valley. This scenario pushes alignment with current French
425 national ecological connectivity strategy to enhance and restore green spaces, connectivity
426 between habitats, biological corridors and biodiversity reservoirs. Despite this expansion of
427 green corridors, and due to the limited spatial extent of such linear features, the two trend
428 scenarios did not significantly alter landscape structure at regional scale, in contrast to the
429 break-away scenarios. Rewilding produced almost total forest colonisation of mountains,
430 while in lowland plains and valleys forest corridors interconnected over time. In the Liberal
431 scenario, the Vercors range remained accessible, and thus relatively attractive for economic

432 activities, which limited landscape changes. In contrast, in the plains and valleys such as
433 Bièvre and Grésivaudan small isolated plots outside large homogenous areas suitable for
434 cereal crops were abandoned and encroached by forest.



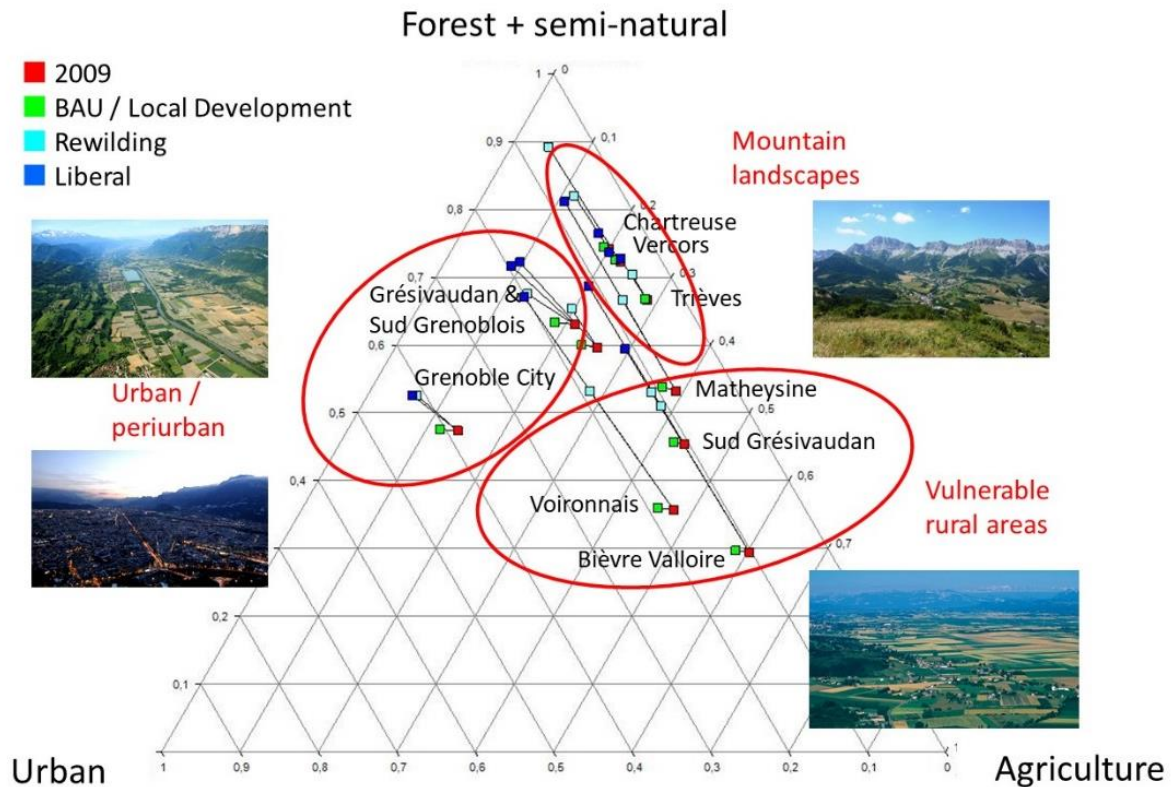
435

436 Figure 3 – Projections of the four scenarios by 2040. (a) proportions of land-use types in
 437 1998, 2009 and for the four scenarios; (b) results of the four scenarios in 2040 over the entire
 438 study area and location of three zoomed areas; (c) zoomed details for the Bièvre, Grésivaudan
 439 and Vercors districts, for 2009 and the four scenarios. Changes between 2009 and 2040

440 projection are highlighted in red (b,c). BAU= Business as usual, LD= Local Development, R=
441 Rewilding, L= Liberal.

442 **3.3. Changes at municipality and sub-regional scale**

443 Results for the municipality-level indicators showed that, initial LULC patterns in 2009 were
444 shaped by the two largest urban centres (Grenoble and Voiron) and the surrounding urban
445 development. Plains harboured predominantly rural municipalities, while forest-dominated
446 municipalities were prevalent in mountains and plateaus (Supplementary figure 2). Under
447 trend scenarios, in periurban municipalities and districts (Grenoble city, Sud Grenoblois,
448 Grésivaudan) urbanisation tracked the 1998-2009 trend, while other districts retained their
449 landscape identity with limited urbanisation except in Voironnais (Figure 4, Supplementary
450 figure 2). Plains and plateaus with initial prevalence of agriculture (45-70% of their total area;
451 Sud Grésivaudan, Voironnais, Bièvre Valloire, Matheysine) were the most sensitive areas to
452 agricultural abandonment and forest recolonization under Rewilding, and even more under the
453 Liberal scenario (with a doubling in forest and semi-natural areas). Scenarios thus showed
454 them to be vulnerable rural areas. In contrast, while forest expansion up to 80-90% of their
455 total area dominated mountain municipalities and districts under Rewilding (Chartreuse,
456 Vercors, Trièves), under the Liberal scenario municipalities in the Vercors and Chartreuse
457 ranges within commuting distance to Grenoble and Voiron retained their rural character with
458 20-30% agricultural land.



459

460 Figure 4 - Aggregated trajectories for districts of the Grenoble urban area. Each square
 461 positions percentage cover in the three-dimension space formed by (1) urban, (2) agriculture
 462 and (3) forest and semi-natural areas for initial state (2009) and the trend (BAU and Local
 463 development were not distinct at this scale), Rewilding and Liberal scenarios. Districts are
 464 clustered (red ellipses) according to their similar initial states and trajectories across scenarios.

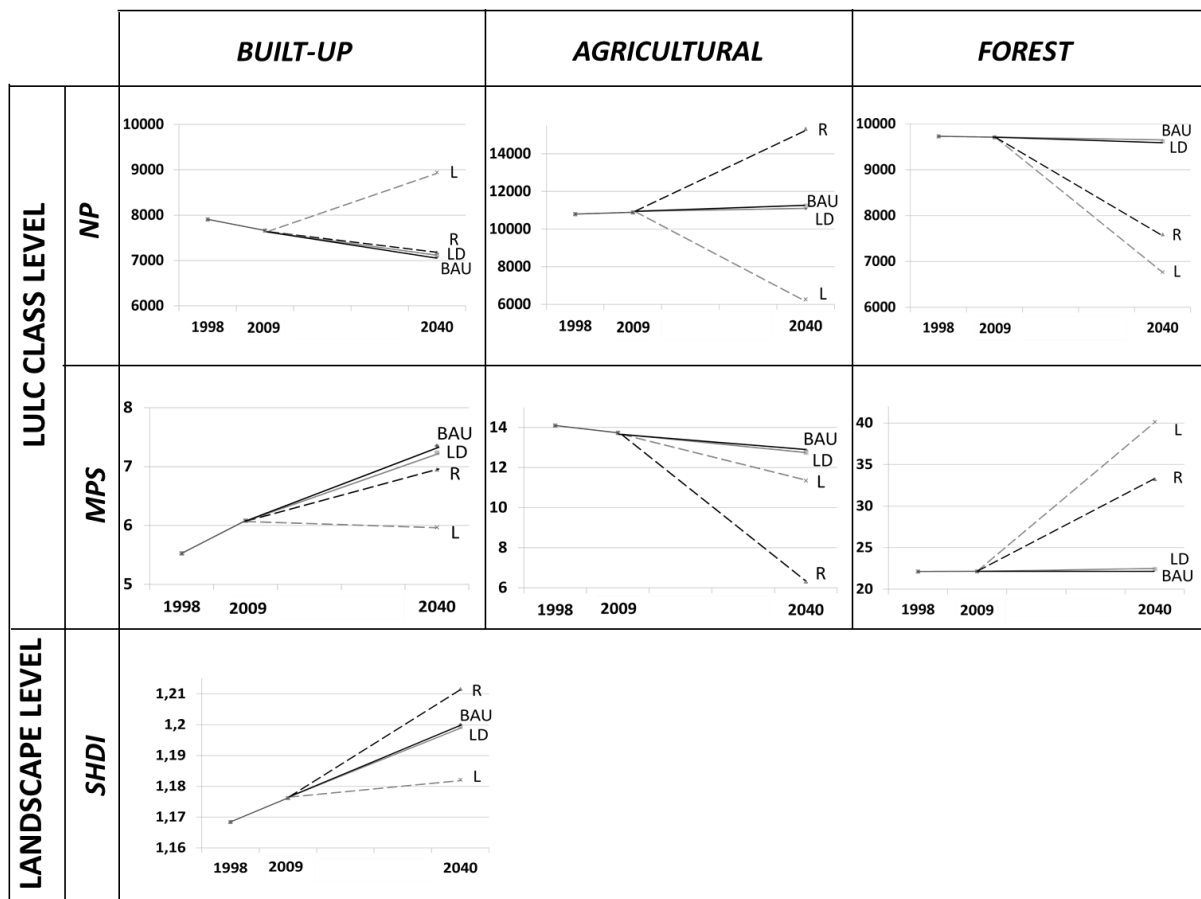
465 3.4. Changes in spatial patterns

466 While model design prescribed consistent mechanisms across scenarios, with urbanisation
 467 occurring at the expense of agricultural land, as did woody encroachment and forest
 468 expansion, loss of agricultural land varied across scenarios, with more or less spatial
 469 continuity, as did the increase in built-up and forested areas. Landscape metrics provided a
 470 finer-scale analysis of these spatial changes within the scenarios. They were complemented by
 471 analyses of ecological connectivity for forest and semi-natural areas (Appendix A).

472 In spite of their slight differences e.g. in green corridor dynamics, the trend scenarios (BAU,
 473 LD) produced similar changes in overall landscape spatial pattern (Figure 5, and
 474 Supplementary table 6 for detailed results). Change rates were unabated from the initial 1998-
 475 2009 period with consolidation into fewer and larger new built-up patches contiguous to
 476 currently existing urban areas (Figure 5). Likewise changes in patch number and size of
 477 individual agricultural and forest land cover types were small and stable over time. Only the
 478 mean size of agricultural patches decreased slightly more in the projections as compared to
 479 1998-2009 trends (while their number remained stable), reflecting consolidation of pre-
 480 existing built-up patches (Figure 5).

481 This contrasts with the two break-away scenarios, with overall much greater changes and
 482 trends not always consistent with those observed between 1998 and 2009 (Figure 5). The two
 483 scenarios were marked by increasing trends in total forested area. These changes of forested
 484 areas are mechanistically linked with those in agricultural land, with the two scenarios
 485 producing opposite changes in spatial patterns: under Rewilding agricultural abandonment
 486 adjacent to existing forest areas increased forest connectivity (see Supplementary analysis 1).
 487 In contrast under the Liberal scenario while abandonment occurred randomly, due to its
 488 magnitude the number and size of agricultural patches decreased, inducing a 30% reduction in
 489 the number of forest patches and a near doubling of forest mean patch size compared to 2009.
 490 The contrasting forest dynamics of the two scenarios were linked with changes in patterns of
 491 built-up land. Under Rewilding urbanisation followed the 1998-2009 trends, with similar
 492 changes in spatial patterns as for the trend scenarios (Figure 5). The Liberal scenario,
 493 however, was marked by an acceleration of peri-urbanisation into agricultural areas with more
 494 numerous and slightly smaller urban patches compared to 2009.

495 The land cover diversity increased slightly between 1998 and 2009, and continued to increase
 496 under the four scenarios (Figure 5 and Supplementary table 6, Shannon Diversity Index
 497 SHDI). While this rate of increase was stable for the two trend scenarios, it increased by half
 498 for the Rewilding scenario due to the predominance of continuous patches of a single LULC
 499 class (forest). Conversely it was halved for the Liberal scenario, reflecting a more even
 500 distribution of LULC classes in a more fragmented landscape (Figure 5).



501

502 Figure 5 – Landscape metrics at the LULC class level (NP: number of patches, MPS: mean
503 patch size) and at landscape level (SDHI: Shannon Diversity Index): columns present LULC
504 metrics for the three main classes undergoing greatest changes. Scenarios: BAU: Business as
505 usual, LD: Local development, R: Rewilding, L: Liberal.

506 **4. Discussion**

507 **4.1. Benefits of participatory normative scenario downscaling**

508 Multi-scale scenarios are considered as particularly relevant to support local or regional
509 decisions by incorporating multiple decision scales, facilitating communication and
510 appropriation by stakeholders and examining local ecological impacts (Biggs et al. 2007).
511 Here we developed a highly participatory downscaling approach allowing a qualitative
512 coupling between normative scenarios designed by policy makers at regional scale and local,
513 spatially explicit dynamics contributed by stakeholders during the participatory process, and
514 refined through quantitative spatial modelling. Four scenarios and accompanying storylines
515 and land use projections translating socio-economic, climate and ecological constraints within
516 normative visions were co-constructed between stakeholders representing main activities and
517 an interdisciplinary research team. Our normative downscaling approach contrasts with
518 common practice for participatory scenario planning (PSP) in place-based socio-ecological
519 research, which has largely favoured exploratory scenarios combining socio-economic and
520 climate drivers (Oteros-Rozas et al. 2015), usually based on bottom-up articulation of past
521 trends and known drivers of land use change and ecosystem service demand (e.g. Hanspach et
522 al. 2014; Mitchell et al. 2015; Schirpke et al. 2017). First, explicit downscaling approaches
523 remain rare in PSP (Harmáčková and Vačkář 2018; Lamarque et al. 2013) probably due to
524 costs and difficulties of such iterative, participatory processes (Walz et al. 2007). While in
525 many PSP processes scenario generation is completed over a short period with a single
526 workshop, here co-production spanned over nearly two years and involved two full time
527 researchers and a team of collaborators contributing the equivalent of another year full time.

528 Second, the lesser adoption of normative scenarios in PSP may be surprising given their value
529 for incorporating stakeholder visions about desirable futures and associated solutions, and for
530 guiding policy and decision-making (Kok et al. 2017). With this study, we contribute to
531 developing practice in normative scenario co-production (Rosa et al. 2017), using an original
532 and replicable participatory downscaling approach combining qualitative and quantitative
533 methods (Harmáčková and Vačkář 2018; Kok et al. 2017; Walz et al. 2007), and that meets
534 criteria of relevance, credibility, legitimacy and creativity (Alcamo et al. 2005).

535 The Montagne 2040 policy initiative, depicting four visions for the Rhône-Alpes region's
536 socio-cultural, economic and governance future was an asset for the project given their high
537 relevance to local policy and planning. Similar to other national or regional initiatives (e.g.
538 Pedroli et al. 2015, Grünfelder et al. 2018), including some national scenarios analysed
539 during our first step scoping (MEDDE 2015), these top-down visions were expressed as main
540 components of socio-economic development for public communication and political action,
541 but without quantification or spatial projections. Familiarity of stakeholders with these initial

542 storylines both facilitated the engagement process for their local adaptation, but also raised
543 normative views and issues of political and power relationships: local stakeholders felt that
544 their innovative (Grenoble was one of the first SCoT plans developed and operationalised in
545 France), and socially and environmentally proactive initiative for reconciling development
546 and conservation of natural capital, was not recognised in Montagne 2040. Stakeholder
547 involvement into adapting storylines insures that their expectations and local context are
548 incorporated, thereby strengthening legitimacy (Castella et al. 2014). As a case in point, the
549 Local Development scenario was developed to address this concern. Although at the time
550 horizon considered here, projected land cover differences were minimal with the Business as
551 usual scenario, researchers' effort for adding this scenario were essential for legitimacy.
552 Stakeholders considered the more extreme scenarios from Montagne 2040 (Rewilding and
553 Liberal) as push-backs, and never fully appropriated Rewilding (Brunet et al. 2018).
554 Nevertheless researchers insisted on developing this scenario, which reflects a political debate
555 in Europe (Pettorelli et al. 2018). Such a give and take attitude is critical for successful
556 transdisciplinary research (Mauser et al. 2013).

557 Given the shared objective between researchers and stakeholders of incorporating ecosystem
558 services into local planning from which the project originated, we needed to translate the
559 directions articulated by Montagne 2040 storylines into land cover maps for subsequent ES
560 modelling (Albert et al. 2014). Downscaling requires careful analysis by researchers of policy,
561 regulation and external scenario documents to specify and quantify expected changes under
562 each vision. Here, guidance from stakeholders was critical for identifying relevant documents
563 and information, along with their specific inputs for missing parameters. Mountain regions
564 require intensive efforts for incorporating their biophysical constraints and associated social
565 contexts into detailed scenarios (Lamarque et al. 2013; Vacquie et al. 2015; Walz et al. 2007).
566 Consistent with other PSP initiatives (Oteros-Rozas et al. 2015), the workshop provided a
567 creative space where stakeholder provided in-depth, spatially-explicit knowledge and
568 imaginative suggestions for the specification of scenarios for the eight sub-regional districts
569 (Brunet et al. 2018). Furthermore, some strongly normative statements were made during this
570 process, especially on power relationships and socio-cultural legacies likely to favour or limit
571 innovation in different districts. This spatial specification was further enriched during the
572 model parameterisation process by joint inputs from stakeholders and local scientific or
573 technical experts. Ultimately, our iterative combination of local stakeholder expertise and
574 planning document analyses, enabled district-specific parameterisation of state-of-the-art
575 LULC models. Credibility and legitimacy of storylines and LULC projections were validated
576 during a next-stage workshop in September 2015, where outputs were presented to the full
577 stakeholder group as an introduction to the participatory analysis of future ecosystem service
578 trade-offs. Main resulting modifications regarded naming and details of some of the more
579 contested storylines, namely Local Development and Rewilding.

580 **4.2. Projecting scenario land use impacts**

581 Consistent with stakeholder expectations and the characteristics of the study area we chose to
582 implement three nested LULC change models for urban, agricultural and forest areas. Each of
583 these models and their scales of implementation were selected according to our analysis of

584 recent dynamics (1998-2009; Vannier et al., 2016) and to data availability (Magliocca et al.,
585 2015). LULC scenario modelling studies in mountains have instead used integrated spatial
586 modelling platforms (FOREcasting SCEnarios - Sohl and Sayler, 2008 ; Land Change
587 Modeler - Eastman, 2012 in the Pyrenees - Vacquie et al. (2015) and Houet et al. (2015);
588 SPA-LUCC in the Austrian Alps - Schirpke et al. 2012), which are more generic and
589 replicable. First rather than combining deterministic (agricultural and forest areas) and
590 probabilistic (urban areas) methods as done here, they rely on common probabilistic models
591 (Magliocca et al., 2015; Sohl and Sayler, 2008; Verburg et al., 2002), which they typically
592 apply to simpler LULC typologies (7 classes on average) across smaller areas (from 7-35 km²
593 - Schirpke et al., 2012, to 498 km² - Houet et al. 2015). Second, these models are
594 parameterised and validated by multi-decadal LULC records (e.g. Tasser et al., 2007 in the
595 Austrian Alps), but are not robust for modelling break-away scenarios.

596 An alternative, more complex and intensive approach was motivated by our multi-scenario
597 objective, and by a search for the necessary spatial and typological precision across a highly
598 diverse and heterogeneous region (Schirpke et al. 2017; Stürck and Verburg 2017). This
599 however implied an enormous parameterisation effort for working at the agricultural parcel
600 scale across an extent of 4450 km², with 41 land cover classes and specific parameters for
601 eight heterogeneous districts. We nevertheless recommend such precision for LULC in
602 heterogeneous, fine-grained landscapes, where processes of urban sprawl, changes in
603 agricultural practices or land abandonment operate at very fine scales and, except for urban
604 conversion, with gradual transitions rather than first-level LULC class conversions, which are
605 relevant for ES modelling (Schirpke et al. 2012, Qiu and Turner 2013, Lasseur et al. 2018).
606 We nevertheless acknowledge that even if pixel-level model allocations are necessarily
607 uncertain as in any LULC model, projections enabled a precise description of changes in
608 landscape patterns and practices at relevant scales for decision makers, namely municipality
609 or district level. Our original LULC maps for the 1998-2009 period had a general mapping
610 precision of at least 95% for level 3 typology (Vannier et al. 2016), and precision for crop
611 successions was typically 35-88% (Lasseur et al. 2018). The spatial precision of the
612 probabilistic model of urban dynamics was estimated to be greater than 10% at pixel level
613 (Longaretti, unpublished data), and by construction the model was implemented so as to
614 exactly reach change targets prescribed for each district. The appropriate scale for use of the
615 maps and their uncertainties were clearly communicated and very well understood by
616 stakeholders during subsequent steps of the work.

617 Projected scenario impacts were consistent with modelling studies for European mountain
618 regions, showing polarisation of landscapes through urbanisation at the expense of
619 agricultural land and forest colonisation of less productive areas (Schirpke et al., 2012; Houet
620 et al. 2015 ; Vacquie et al., 2015; Stürck et al. 2016). However, the scenarios produced
621 contrasting spatial patterns. While the two trend scenarios showed typical European patterns
622 of spatially-continuous urban expansion into agricultural land (Stürck et al. 2016), the two
623 break-away scenarios resulted in strong contrasts with 2009, and amongst themselves due to
624 spatial contiguous vs. random land abandonment and reforestation. The reforestation of less
625 productive land and the resulting landscape homogenisation under liberal economic settings is

626 a common feature of scenarios for mountains (Schirpke et al., 2012; Vacquie et al., 2015;
627 Brunner et al. 2017) and other cultural landscapes (Hanspach et al. 2014; Plieninger et al.
628 2013), and at European scale (Stürck et al. 2016). However the deliberately contiguous
629 pattern proposed under Rewilding for developing ecological connectivity has rarely been
630 considered in spite of this scenario's plausibility in the European policy context (Schulp et al.
631 2016; Stürck et al. 2016) and growing interest by the conservation community (Pettorelli et al.
632 2018). Landscapes metrics strongly benefit land planning in addition to analyses of change
633 volumes (De Vreese et al., 2016), especially when applied to scenarios (Lausch et al., 2015).
634 Given European and national green and blue corridors policy targets, it is essential to
635 document alternatives in terms of landscape pattern and connectivity (De Vreese et al., 2016).
636 Connectivity analysis also integrates relevant ecological characteristics (Rao et al., 2019). The
637 value of such analyses was thus evident for distinguishing environmental benefits across the
638 two break-away scenarios. On the other hand, while stakeholders insisted in distinguishing the
639 Local development scenarios from Business as Usual based on governance and stronger urban
640 consolidation constraints, spatial differences were not detectable. We expect that, given the
641 relatively low rates of urban expansion, their differences in urban growth forms would
642 become evident over longer time horizons. Lastly, connectivity in agricultural areas improved
643 under all scenarios, complying with European and national legislation (French Law for
644 Biodiversity and Landscapes 2016).

645 **4.3. Implications for ecosystem services**

646 The use of scenarios offers new perspectives for integrated planning that takes into account
647 ecological dynamics and ecosystem services (Opdam et al. 2015). Significant implications of
648 each scenario and associated LULC projections for future ecosystem service supply capacity
649 are expected. Apart from obvious differences in provisioning services across scenarios due to
650 their fundamentally different economies, projected changes in land cover would differently
651 impact regulation services that strongly depend on forest cover such as carbon storage, water
652 quality and quantity regulation or erosion and rockfall control. While increased wood stocks
653 in the two break-away scenarios would increase carbon storage, their economic context would
654 not necessarily promote wood production (Lafond et al. 2017). Their positive effects on
655 regulation services would also trade-off with loss in crop and fodder provisioning
656 (Harmáčková and Vačkář 2018; Schirpke et al. 2017; Stürck and Verburg 2017). Scenario
657 contrasts in forest cover and spatial pattern, agricultural land and urban development would
658 also affect cultural services as limited forest expansion is perceived positively (e.g. recreation,
659 Byczek et al. 2018; aesthetic value – Schirpke et al. 2019) and favours some protected
660 species. Spatial differences between scenarios will specifically impact regulation services
661 dependent on lateral flows of matter and organisms (e.g. water quality and quantity
662 regulation, erosion control, pollination; Verhagen et al. 2016) or cultural services depending
663 on landscape connectivity (e.g. cultural value of protected vertebrates; Schirpke et al. 2018) or
664 landscape heterogeneity (e.g. outdoor recreation; Byczek et al. 2018).

665 **5. Concluding remarks: Implications for land use planning and** 666 **decision**

667 Through a structured and sustained two-year participatory process, our interdisciplinary
668 research team co-produced with local stakeholders scenario narratives and associated land use
669 projections downscaling four normative scenarios produced by the administrative region's
670 government. This process relevant to similar urban regions in developed mountain and other
671 regions fostered (i) local appropriation of top-down visions, (ii) incorporation of participants
672 normative views, (iii) simultaneous consideration of local initiatives for reconciling economic
673 development with the conservation of natural resources and processes, and of national and
674 European policy challenges, and (iv) incorporation of biophysical and socio-economic
675 heterogeneity and legacies. Final mapped scenarios described how landscape transformations
676 that are common across mountain and other culturally valued regions would unfold in the
677 Grenoble context. They highlighted how pairs of scenarios distinct in their baseline values
678 and associated governance, namely the two trend scenarios (BAU and Local development) or
679 the two break-away scenarios (Rewilding and Liberal), could converge to similar landscape
680 outcomes – curbing periurban sprawl or extensive forest expansion respectively.
681 Nevertheless, the stark contrast in landscape patterns for the two break-away scenarios
682 strongly supported the use of a fine-scale, detailed spatially-explicit approach incorporating
683 sub-regional specificities essential to stakeholders. As such projected LULC maps, along with
684 their detailed context elements and parameters, can readily be used by land planners and
685 nature managers. For instance, they are of direct relevance for the ongoing implementation of
686 the French national ecological connectivity strategy, or for the management and development
687 of natural protected areas – including a new regional park proposed for the Belledonne range.
688 Forthcoming projections of scenario impacts on current bundles of ecosystem services
689 (Vannier et al., 2019) will add to land planners and decision managers baseline knowledge
690 and know how, and challenge their preconceptions of the costs and benefits of alternative
691 development trajectories (Brunet et al. 2018).

692 **References**

- 693 Albert C, Aronson J, Fürst C, Opdam P (2014) Integrating ecosystem services in landscape
694 planning: requirements, approaches, and impacts. *Landscape Ecology* 29(8):1277-
695 1285
- 696 Alcamo J (ed) (2009) *Environmental Futures: The Practice of Environmental Scenario*
697 *Analysis*. Elsevier
- 698 Alcamo J, van Vuuren D, Ringler C et al (2005) Changes in nature's balance sheet: model-
699 based estimates of future worldwide ecosystems. *Ecology and Society* 10(2):19
- 700 Bierry, A., Lavorel, S., 2016. Implication des parties prenantes d'un projet de territoire dans
701 l'élaboration d'une recherche à visée opérationnelle. *Sciences, Eaux & Territoires* 21,
702 [http://www.set-revue.fr/sites/default/files/articles/pdf/set-revue-gestion-territoires-](http://www.set-revue.fr/sites/default/files/articles/pdf/set-revue-gestion-territoires-recherche-implication-acteurs.pdf)
703 [recherche-implication-acteurs.pdf](http://www.set-revue.fr/sites/default/files/articles/pdf/set-revue-gestion-territoires-recherche-implication-acteurs.pdf).
- 704 Biggs R, Raudsepp-Hearne C, Atkinson-Palombo C et al (2007) Linking Futures across
705 Scales: a Dialog on Multiscale Scenarios. *Ecology and Society* 12(1)

706 Bohunovsky L, Jäger J, Omann I (2011) Participatory scenario development for integrated
707 sustainability assessment. *Regional Environmental Change* 11(2):271-284

708 Brunet L, Tuomisaari J, Lavorel S et al (2018) Actionable knowledge for land-use planning:
709 making ecosystem services operational. *Land Use and Policy* 72:27-34

710 Brunner SH, Huber R, Grêt-Regamey A (2017) Mapping uncertainties in the future provision
711 of ecosystem services in a mountain region in Switzerland. *Regional Environmental*
712 *Change*

713 Byczek, C., Longaretti, P.-Y., Renaud, J., Lavorel, S. (2018) Benefits of crowd-sourced GPS
714 information for modelling the recreation ecosystem service. *PLOS ONE* 13,
715 e0202645. <https://doi.org/10.1371/journal.pone.0202645>

716 Cabral P, Feger C, Levrel H, Chambolle M, Basque D (2016) Assessing the impact of land-
717 cover changes on ecosystem services: A first step toward integrative planning in
718 Bordeaux, France. *Ecosystem Services* 22, part B:318-327

719 Carpenter SR, Mooney HA, Agard J et al (2009) Science for managing ecosystem services:
720 Beyond the Millennium Ecosystem Assessment. *Proceedings of the National*
721 *Academy of Sciences* 106:1305-1312

722 Castella, J.-C., J. Bourgoïn, G. Lestrelin and B. Bouahom (2014). "A model of the science–
723 practice–policy interface in participatory land-use planning: lessons from Laos."
724 *Landscape Ecology* 29(6): 1095-1107.

725 Centre Economique, Social et Environnemental Régional Rhône-Alpes (2013) *Montagne*
726 *2040*. 2013-03. Région Rhône-Alpes, pp. 228

727 Colloff MJ, Martín-López B, Lavorel S et al (2017) An integrative framework for enabling
728 transformative adaptation. *Environmental Science & Policy* 68:87-96

729 Cradock-Henry, N.A., Frame, B., Preston, B.L., Reisinger, A., Rothman, D.S., 2018.
730 Dynamic adaptive pathways in downscaled climate change scenarios. *Climatic Change*
731 150, 333-341.

732 Cushman, S.A., McGarigal, K., Neel, M.C., 2008. Parsimony in landscape metrics: Strength,
733 universality, and consistency. *Ecol. Indic.* 8, 691–703.
734 <https://doi.org/10.1016/j.ecolind.2007.12.002>

735 De Vreese, R., Leys, M., Fontaine, C.M., Dendoncker, N., 2016. Social mapping of perceived
736 ecosystem services supply – The role of social landscape metrics and social hotspots
737 for integrated ecosystem services assessment, landscape planning and management.
738 *Ecological Indicators* 66, 517–533. <https://doi.org/10.1016/j.ecolind.2016.01.048>

739 Díaz S, Demissew S, Carabias J et al (2015) The IPBES Conceptual Framework —
740 connecting nature and people. *Current Opinion in Environmental Sustainability*
741 14(0):1-16

742 Eastman, 2012. *Idrisi Selva, Guide to GIS and Image Processing*.

743 Griggs D, Stafford-Smith M, Gaffney O et al (2013) Sustainable development goals for
744 people and planet. *Nature* 495:305

745 Grünenfelder, P., Schellenbauer, P., Dümmler, P., Langenegger, J., Parzer-Epp, V., Salvi, M.,
746 Schaad, J., Schnell, F., Steiner, U., 2018. *Livre Blanc Suisse - Six esquisses pour*
747 *l'avenir*. *Avenir Suisse, Zürich*, p. 35.

748 Hanspach J, Hartel T, Milcu AI et al (2014) A holistic approach to studying social-ecological
749 systems and its application to southern Transylvania. *Ecology and Society* 19(4)

750 Harmáčková ZV, Vačkář D (2018) Future uncertainty in scenarios of ecosystem services
751 provision: Linking differences among narratives and outcomes. *Ecosystem Services*
752 Harrison PA, Dunford R, Savin C et al (2015) Cross-sectoral impacts of climate change and
753 socio-economic change for multiple, European land- and water-based sectors. *Climatic*
754 *Change* 128:279-294

755 Houet, T., Vacquié, L., Sheeren, D. (2015) Evaluating the spatial uncertainty of future land
756 abandonment in a mountain valley (Vicdessos, Pyrenees - France): Insights from
757 model parameterization and experiments. *J. Mt. Sci.* 12, 1095–1112.
758 <https://doi.org/10.1007/s11629-014-3404-7>

759 Kabisch N, Frantzeskaki N, Pauleit S et al (2016) Nature-based solutions to climate change
760 mitigation and adaptation in urban areas: perspectives on indicators, knowledge gaps,
761 barriers, and opportunities for action. *Ecology and Society* 21(2)

762 Kohler M, Stotten R, Steinbacher M et al (2017) Participative Spatial Scenario Analysis for
763 Alpine Ecosystems. *Environmental Management* 60(4):679-692

764 Kok K, Biggs R, Zurek M (2007) Methods for developing multiscale participatory scenarios:
765 insights from southern Africa and Europe. *Ecology and Society* 13

766 Kok MTJ, Kok K, Peterson GD, Hill R, Agard J, Carpenter SR (2017) Biodiversity and
767 ecosystem services require IPBES to take novel approach to scenarios. *Sustainability*
768 *Science* 12(1):177-181

769 Lafond, V., T. Cordonnier, Z. Mao and B. Courbaud (2017). "Trade-offs and synergies
770 between ecosystem services in uneven-aged mountain forests: evidences using Pareto
771 fronts." *European Journal of Forest Research*: 1-16.

772 Lamarque P, Artaux A, Nettié B, Dobremez L, Barnaud C, Lavorel S (2013) Taking into
773 account farmers' decision making to map fine-scale land management adaptation to
774 climate and socio-economic scenarios. *Landscape and Urban Planning* 119:147-157

775 Lasseur, R., Vannier, C., Lefebvre, J., Longaretti, P.-Y., Lavorel, S. (2018) Landscape-scale
776 modeling of agricultural land use for the quantification of ecosystem services. *J. Appl.*
777 *Remote Sens.* 12, 046024. <https://doi.org/10.1117/1.JRS.12.046024>

778 Lausch, A., Blaschke, T., Haase, D., Herzog, F., Syrbe, R.-U., Tischendorf, L., Walz, U.,
779 2015. Understanding and quantifying landscape structure – A review on relevant
780 process characteristics, data models and landscape metrics. *Ecological Modelling, Use*
781 *of ecological indicators in models* 295, 31–41.
782 <https://doi.org/10.1016/j.ecolmodel.2014.08.018>

783 Lavorel, S., Colloff, M.J., Locatelli, B., Gorrdard, R., Prober, S.M., Gabillet, M., Devaux, C.,
784 Laforgue, D., Peyrache-Gadeau, V., 2019. Mustering the power of ecosystems for
785 adaptation to climate change. *Environ. Sci. Policy* 92, 87–97.
786 <https://doi.org/10.1016/j.envsci.2018.11.010>

787 Maes J, Sanders NJ (2017) Nature-based solutions for Europe's sustainable development.
788 *Conservation Letters* 10:121-124

789 Magliocca, N.R., van Vliet, J., Brown, C., Evans, T.P., Houet, T., et al. (2015) From meta-
790 studies to modeling: Using synthesis knowledge to build broadly applicable process-
791 based land change models. *Environ. Model. Softw.* 72, 10–20.
792 <https://doi.org/10.1016/j.envsoft.2015.06.009>

793 Mauser, W., Klepper, G., Rice, M., Schmalzbauer, B.S., Hackmann, H., Leemans, R., Moore,
794 H., 2013. Transdisciplinary global change research: the co-creation of knowledge for
795 sustainability. *Current Opinion in Environmental Sustainability* 5, 420-431.

796 McGarigal, K., Cushman, S., Ene, E. (2012) Spatial Pattern Analysis Program for Categorical
797 and Continuous Maps. Computer software program produced by the authors at the
798 University of Massachusetts, Amherst. Available at the following web site:
799 <http://www.umass.edu/landeco/research/fragstats/fragstats.html>.

800 Ministère de l'Écologie, du Développement Durable et de l'Énergie, 2015, Territoire Durable
801 2030, commissariat général au développement durable, mission prospective.
802 <http://www.territoire-durable-2030.developpement-durable.gouv.fr/>

803 Mitchell MGE, Bennett EM, Gonzalez A et al (2015) The Montérégie Connection: linking
804 landscapes, biodiversity, and ecosystem services to improve decision making. *Ecology
805 and Society* 20(4)

806 Moss RH, Edmonds JA, Hibbard KA et al (2010) The next generation of scenarios for climate
807 change research and assessment. *Nature* 463(7282):747-756

808 Nesshöver C, Assmuth T, Irvine KN et al (2017) The science, policy and practice of nature-
809 based solutions: An interdisciplinary perspective. *Science of The Total Environment*
810 579:1215-1227

811 Nieto-Romero M, Milcu A, Leventon J, Mikulcak F, Fischer J (2016) The role of scenarios in
812 fostering collective action for sustainable development: Lessons from central
813 Romania. *Land Use Policy* 50:156-168

814 Opdam P, Coninx I, Dewulf A, Steingröver E, Vos C, van der Wal M (2015) Framing
815 ecosystem services: Affecting behaviour of actors in collaborative landscape
816 planning? *Land Use Policy* 46(0):223-231

817 Oteros-Rozas E, Martín-López B, Daw T et al (2015) Participatory scenario-planning in
818 place-based social-ecological research: insights and experiences from 23 case studies.
819 *Ecology & Society* 20(4):32

820 Oteros-Rozas E, Martín-López B, López CA, Palomo I, González JA (2013) Envisioning the
821 future of transhumant pastoralism through participatory scenario planning: a case
822 study in Spain. *The Rangeland Journal* 35(3):251-272

823 Palomo I, Martín-López B, López-Santiago C, Montes C (2011) Participatory Scenario
824 Planning for Protected Areas Management under the Ecosystem Services Framework:
825 the Donana Social-Ecological System in Southwestern Spain. *Ecology and Society*
826 16(1)

827 Pedrolí, B., Rounsevell, M. D. A., Metzger, M. J., Paterson, J., & The VOLANTE
828 Consortium. (2015). *The VOLANTE Roadmap towards sustainable land resource
829 management in Europe. VOLANTE final project document*. Wageningen: Alterra
830 Wageningen UR.

831 Peterson GD, Cumming GS, Carpenter SR (2003) Scenario Planning: a Tool for Conservation
832 in an Uncertain World

833 Pettorelli, N., Barlow, J., Stephens, P.A., Durant, S.M., Connor, B., Schulte to Bühne, H.,
834 Sandom, C.J., Wentworth, J., Toit, J.T., 2018. Making rewilding fit for policy. *Journal
835 of Applied Ecology* 55, 1114-1125.

836 Plieninger T, Bieling C, Ohnesorge B, Schaich H, Schleyer C, Wolff F (2013) Exploring
837 Futures of Ecosystem Services in Cultural Landscapes through Participatory Scenario
838 Development in the Swabian Alb, Germany. *Ecology and Society* 18(3)

839 Plieninger T, van der Horst D, Schleyer C, Bieling C (2014) Sustaining ecosystem services in
840 cultural landscapes. *Ecology and Society* 19(2)

841 Qiu, J., Turner, M.G., 2013. Spatial interactions among ecosystem services in an urbanizing
842 agricultural watershed. *Proceedings of the National Academy of Sciences* 110, 12149-
843 12154.

844 Rao, Y., Zhang, J., Wang, K., Wu, X., 2019. How to prioritize protected areas: A novel
845 perspective using multidimensional land use characteristics. *Land Use Policy* 83, 1–
846 12. <https://doi.org/10.1016/j.landusepol.2019.01.023>

847 Reed MS, Hubacek K, Bonn A et al (2013) Anticipating and Managing Future Trade-offs and
848 Complementarities between Ecosystem Services. *Ecology and Society* 18(1)

849 Rosa IMD, Pereira HM, Ferrier S et al (2017) Multiscale scenarios for nature futures. *Nature*
850 *Ecology & Evolution* 1(10):1416-1419

851 Rounsevell MDA, Pedrolì B, Erb K-H et al (2012) Challenges for Land System Science. *Land*
852 *Use and Policy* 29:899-910

853 Ruget F., Bernard F., Durand J.L., Graux A.I., Lacroix B., Moreau J.-C., Ripoche D., 2013, «
854 Impacts des changements climatiques sur les productions de fourrages (prairies,
855 luzerne, maïs) : variabilité selon les régions et les saisons », *Fourrages*, 214, p. 99-110.

856 Schirpke U, Kohler M, Leitinger G, Fontana V, Tasser E, Tappeiner U (2017) Future impacts
857 of changing land-use and climate on ecosystem services and resilience of mountain
858 grassland. *Ecosystem Services* 26:79-94

859 Schirpke, U., Leitinger, G., Tappeiner, U., Tasser, E., 2012. SPA-LUCC: Developing land-
860 use/cover scenarios in mountain landscapes. *Ecol. Inform.* 12, 68–76.
861 <https://doi.org/10.1016/j.ecoinf.2012.09.002>

862 Schirpke, U., Meisch, C., Tappeiner, U., 2018. Symbolic species as a cultural ecosystem
863 service in the European Alps: insights and open issues. *Landscape Ecology* 33, 711-
864 730.

865 Schirpke, U., G. Tappeiner, E. Tasser and U. Tappeiner (2019). "Using conjoint analysis to
866 gain deeper insights into aesthetic landscape preferences." *Ecological Indicators* 96:
867 202-212.

868 Schulp CJE, Van Teeffelen AJA, Tucker G, Verburg PH (2016) A quantitative assessment of
869 policy options for no net loss of biodiversity and ecosystem services in the European
870 Union. *Land Use Policy* 57:151-163

871 Seidl, R., Schelhaas, M.-J., Lexer, M.J., 2011. Unraveling the drivers of intensifying forest
872 disturbance regimes in Europe. *Glob. Change Biol.* 17, 2842–2852.
873 <https://doi.org/10.1111/j.1365-2486.2011.02452.x>

874 Sharpe, B., Hodgson, A., Leicester, G., Lyon, A., Fazey, I., 2016. Three horizons: a pathways
875 practice for transformation. *Ecol. Soc.* 21. <https://doi.org/10.5751/ES-08388-210247>

876 Steffen W, Richardson K, Rockström J et al (2015) Planetary boundaries: Guiding human
877 development on a changing planet. *Science* 347(6223)

878 Soares-Filho, B., Rogrigues H., Follador, M. (2013) A hybrid analytical-heuristic method for
879 calibrating land use change models, *Environmental Modelling and Software*, 43, 80-
880 87. <https://doi.org/10.1016/j.envsoft.2013.01.010>

- 881 Sohl, T., Sayler, K. (2008) Using the FORE-SCE model to project land-cover change in the
882 southeastern United States. *Ecol. Model.* 219, 17.
883 <https://doi.org/10.1016/j.ecolmodel.2008.08.003>
- 884 Stürck J, Levers C, van der Zanden EH et al (2016) Simulating and delineating future land
885 change trajectories across Europe. *Regional Environmental Change* in press
- 886 Stürck J, Verburg PH (2017) Multifunctionality at what scale? A landscape multifunctionality
887 assessment for the European Union under conditions of land use change. *Landscape*
888 *Ecology* 32:481-500
- 889 Tasser E, Walde J, Tappeiner U, Teutsch A, Nogglner W (2007) Land-use changes and natural
890 reforestation in the Eastern Central Alps. *Agriculture, Ecosystems & Environment*
891 118(1-4):115-129
- 892 Turkelboom F, Leone M, Jacobs S et al (2017) When we cannot have it all: Ecosystem
893 services trade-offs in the context of spatial planning. *Ecosystem Services*:in press
- 894 Vacquie L, Houet T, Sohl T, Reker R, Sayler K (2015) Modelling regional land change
895 scenarios to assess land abandonment and reforestation dynamics in the Pyrenees
896 (France). *J. Mt. Sci.* 12(4):905-920
- 897 Van Kerkhoff, L., Munera, C., Dudley, N., Guevara, O., Wyborn, C., Figueroa, C., Dunlop,
898 M., Hoyos, M.A., Castiblanco, J., Becerra, L., 2018. Towards future-oriented
899 conservation: Managing protected areas in an era of climate change. *Ambio*.
- 900 Vannier C., Lasseur R., Crouzat E., Byczek C., Lafond V., Cordonnier T., Longaretti P.Y.,
901 Lavorel S. (2019) Mapping Ecosystem Services bundles in a heterogeneous mountain
902 region, *Ecosystems and People*, 15(1), 74-88.
903 <https://doi.org/10.1080/26395916.2019.1570971>
- 904 Vannier, C., Lefebvre, J., Longaretti, P.-Y., Lavorel, S. (2016) Patterns of landscape change
905 in a rapidly urbanizing mountain region. *Cybergeo Eur. J. Geogr.*
906 <https://doi.org/10.4000/cybergeo.27800>
- 907 Verburg, P.H., Soepboer, W., Veldkamp, A., Limpiada, R., Espaldon, V., Mastura, S.S.A.,
908 2002. Modeling the Spatial Dynamics of Regional Land Use: The CLUE-S Model.
909 *Environ. Manage.* 30, 391–405. <https://doi.org/10.1007/s00267-002-2630-x>
- 910 Verhagen, W., Van Teeffelen, A.J.A., Baggio Compagnucci, A., Poggio, L., Gimona, A.,
911 Verburg, P.H., 2016. Effects of landscape configuration on mapping ecosystem
912 service capacity: a review of evidence and a case study in Scotland. *Landscape*
913 *Ecology* 41, 1457-1479.
- 914 Verkerk, P.J., Lindner, M., Pérez-Soba, M., Paterson, J.S., Helming, J., Verburg, P.H.,
915 Kuemmerle, T., Lotze-Campen, H., Moiseyev, A., Müller, D., Popp, A., Schulp,
916 C.J.E., Stürck, J., Tabeau, A., Wolfslehner, B., van der Zanden, E.H., 2018.
917 Identifying pathways to visions of future land use in Europe. *Regional Environmental*
918 *Change* 18, 817-830.
- 919 Walz A, Lardelli C, Behrendt H et al (2007) Participatory scenario analysis for integrated
920 regional modelling. *Landscape and Urban Planning* 81(1):114-131