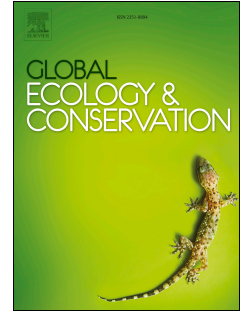


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Using key biodiversity areas to guide effective expansion of the global protected area network

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1 **Using Key Biodiversity Areas to guide effective expansion of the**
2 **global protected area network**

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23

24 **Abstract**

25 Using spatial prioritization, we identify priority areas for the expansion of the global
26 protected area network. We identify a set of unprotected key biodiversity areas
27 (KBAs) that would efficiently complement the current protected area network in
28 terms of coverage of ranges of terrestrial vertebrates. We show that protecting a
29 small fraction (0.36%) of terrestrial area within KBAs could increase conservation
30 coverage of ranges of threatened vertebrates by on average 14.7 percentage points.
31 We also identify areas outside both the protected area and KBA networks that
32 would further complement the priority KBAs. These areas are likely to hold
33 populations of species that are poorly protected or covered by KBAs, and where on-
34 the-ground surveys might confirm suitability for KBA designation or protection.

35 **1. Introduction**

36 Protected areas (PA) are the cornerstone for halting the global biodiversity crisis
37 (UNEP-CBD, 2010). While there has been a steady increase in the coverage of PAs
38 over the last decades (UNEP-WCMC and IUCN, 2016), further expansion is needed
39 urgently (Tittensor et al., 2014; WWF, 2016). For example, Aichi Target 11 of the
40 Convention on Biological Diversity recommends increasing terrestrial PA coverage
41 to 17% by 2020 (UNEP-CBD, 2010) from current 14.7% (UNEP-WCMC and IUCN,
42 2016). As the need to act is urgent and resources are limited, prioritization of
43 conservation effort is important (McCarthy et al., 2012; Pouzols et al., 2014).

44

45 Global conservation priority rankings have been developed using a variety of
46 methods ranging from simple species richness ranking (Jenkins et al., 2013) to more
47 complex methods that, in addition to biodiversity, also account for additional factors
48 such as costs or land-use change (Butchart et al., 2015; Pouzols et al., 2014; Venter
49 et al., 2014). The sizes of analysis units have ranged from ecoregions down to 1 km²
50 grid cells (Hoekstra et al., 2005; Myers et al., 2000; Pouzols et al., 2014). Although
51 computation power has increased in recent years allowing finer scale analyses also
52 at the global extent, prioritization analyses are still typically being run using rather
53 large grids (from 10km² to 200 km²)(Di Marco et al., 2017). One reason for this is
54 the coarse resolution of globally available datasets - especially the species range
55 maps that most of the analyses are relying on (Di Marco et al., 2017).

56 Global conservation prioritization analyses are needed to identify broad scale
57 conservation priority patterns, to establish general principles about how the global
58 PA network should be expanded, and to estimate how well it covers biodiversity
59 (see for example Butchart et al., 2015; Pouzols et al., 2014; Venter et al., 2014).
60 Nevertheless, the often implicitly assumed link between global conservation
61 priorities and on the ground conservation action is not well established nor
62 discussed widely in most the global extent prioritization analyses (see for example
63 Butchart et al., 2015; Pouzols et al., 2014; Venter et al., 2014). In fact, country level
64 studies have shown mixed results about whether the priority schemes affect the
65 amount of conservation funds directed to the priority areas (Halpern et al., 2006;
66 Holmes et al., 2012).

67

68 One reason hindering the usage of global conservation prioritization analyses to
69 inform actual planning could be that it is hard to draw concrete suggestions for local
70 actions based on coarse scale or vaguely delineated global priority patterns. On
71 average PAs are much smaller than the planning units typically used in global scale
72 conservation prioritization analyses. Usage of large grids as planning units can also
73 lead to inefficiency in the prioritization analyses (Di Marco et al., 2017). On the other
74 hand, inevitable lack of detail about the local circumstances in the global analyses
75 might hinder the usage of the finer resolution global analyses to guide planning at
76 local level.

77

78 Key Biodiversity Areas (KBA) are promoted by the IUCN as a means to identify "sites
79 of importance for the global persistence of biodiversity" (IUCN, 2016a). KBAs are
80 established based on clearly defined rules against which individual sites are
81 matched (IUCN, 2016a). In contrast to global conservation priority analyses that
82 usually consider all areas simultaneously and require spatial data across the full
83 study area, the KBA method is applied site by site using locally available data (IUCN,
84 2016a). Focusing on one site at time makes it possible to use, or even collect,
85 detailed information that is needed for delineating conservation areas in a way that
86 accounts for e.g. local ecological processes or socio-economic reality. On the other
87 hand, being a local, site-based approach, KBAs cannot directly account for network-
88 level factors, such as balance between different species (complementarity) or
89 representativeness of the network as a whole (Moilanen et al., 2009). This could
90 potentially lead to a situation where most of the available resources are directed to
91 areas having similar species composition, or species that might already be well
92 covered elsewhere in the protected area network, while some other species might
93 be completely missing.

94

95 According to IUCN standards, KBAs should be delineated so that they are
96 manageable units, accounting for local ecological, physical and socio-economic
97 contexts (IUCN, 2016a). These factors are important for management of any

98 conservation areas, which makes the KBA delineation information, that is based on
99 detailed local level information but available globally, a valuable data resource for
100 global conservation prioritization analyses. Increasing KBA protection is generally
101 considered to be critical to enhancing species persistence (Butchart et al., 2015,
102 2012). Indeed, one of the five main indicators of progress towards the Aichi target
103 11 is PA coverage of KBAs (UNEP-WCMC, 2017). Nevertheless, only one-fifth of the
104 KBAs are reported to be fully protected (Butchart et al., 2015). Hence, unprotected
105 KBAs are prime candidates for global PA network expansion (Butchart et al., 2015,
106 2012).

107

108 In this paper we explore the expansion of the global PA network by combining KBAs
109 as manageable conservation units with the ability of conservation prioritization
110 software to find globally effective solutions. Using the Zonation software (Moilanen
111 et al. 2014), we identify global conservation priority areas for expansion of the PA
112 network by highlighting a set of unprotected KBAs that, if protected, would area-
113 efficiently increase mean coverage of threatened vertebrate species ranges in the
114 global PA network while improving balance by paying highest attention to species
115 with lowest coverage. We further identify the priority areas with most urgent need
116 for action by considering the human influence index (HII) within priority sites. To
117 reduce effects of uncertainty associated with range maps, we used species
118 observations from GBIF (Global Biodiversity Information Facility, GBIF 2017) to up

119 weight areas with confirmed sightings. By using KBAs as planning units, we aimed
120 to reduce effects of data uncertainties and overcome some of the limitations that
121 follow from identifying conservation priorities based on large unmanageable areas
122 or pixels that are too small to capture ecological processes (Hurlbert & Jetz, 2007).
123 Our view is that using KBAs as planning units can also help bridge the gap between
124 global conservation priority analyses and site level conservation action, as KBAs are
125 sites that could well be the focus of immediate protection. We also identify priority
126 areas outside the PA and KBA networks that would further complement the priority
127 KBAs. These areas are identified using a grid based analysis and thus, compared to
128 the priority KBAs, require more information to confirm their suitability for
129 conservation.

130

131 **2. Methods**

132

133 **2.1. Data manipulation**

134 We rasterized all spatial data using the intersect method, geographic coordinate
135 system and 1 arc-minute (equal to 1.85 km at the equator) resolution. Such fine
136 resolution was needed to approximate the location and shape of PAs and KBAs with
137 reasonable accuracy. This raster resolution should not be confused with the size of
138 planning units in spatial analysis, which in our main analysis was determined by the

139 size of the unprotected KBAs. All data processing was performed with R v. 3.4.1 (R
140 Core Team, 2017). Latitudinal variation in cell size was accounted for in Zonation
141 analyses and data processing.

142

143 **2.2. Species range maps**

144 To identify priority KBAs, we used range maps of threatened (Critically Endangered,
145 Endangered or Vulnerable) terrestrial mammals, birds and amphibians (n = 4,892
146 species) in the IUCN Red List (IUCN, 2016b). In the priority analysis of areas outside
147 the KBA network we also accounted for Data Deficient species (n = 2,336 species).

148 For bird species with different seasonal ranges (e.g. *Acrocephalus paludicola*,
149 *Emberiza aureola*), we included all ranges as separate feature layers. This promotes
150 equal coverage of all areas that are important for the survival of migratory species.
151 Combining wintering and breeding ranges to a single input layer could lead to a
152 situation where a species is considered to be well covered by the prioritization, but
153 would totally lack either wintering or breeding range, potentially affecting species
154 long term survival. Using separate input layers forces Zonation to seek balance
155 between wintering and breeding ranges and to account for both of them in the
156 priority areas.

157

158 **2.3. Protected area and KBA data**

159 PA data was extracted from the World Database on Protected Areas (IUCN & UNEP-
160 WCMC, 2016) and KBA data from the World Database of Key Biodiversity Areas
161 (KBA Partnership, 2016). We included all designated terrestrial PAs and KBAs that
162 had polygonal representation. PAs represented only by points were discarded,
163 because without accurate information, we thought it safer to underestimate than to
164 overestimate PA coverage (see Visconti et al. 2013 for further discussion). The
165 original KBA polygon data set included approximately 14,900 KBAs that were
166 reduced to 13,700 after rasterization and removal of marine areas. Unprotected
167 KBAs were then identified by overlaying the KBA and PA rasters. The KBA data set
168 also included information about the criteria according to which the site was
169 assigned the KBA status, We used this information to explore whether our
170 prioritization method would give higher priority to KBAs that were established due
171 to occurrence of threatened species as compared to other criteria triggering KBA
172 definition. This could be expected because we used the range maps of threatened
173 species as primary data in the analysis.

174

175 **2.4. Species observations**

176 Conservation prioritization analyses that use species range maps as an input data
177 are prone to commission errors, in which species are erroneously thought to be
178 present where it actually is not (Rondinini et al. 2006). We believe that using KBAs

179 as planning units can reduce commission errors, because KBAs have been shown to
180 harbor more threatened species than their surroundings (Di Marco et al. 2016). To
181 reduce commission errors also in the analysis focusing on areas outside the KBA
182 network, we decided to use GBIF observation as additional information about
183 species occurrence. The rationale behind this approach is that the areas of the
184 species ranges where the species has also been observed are also more likely to
185 actually harbor the species. We acknowledge that this method alone does not solve
186 the problem of false positives, but could contribute towards identifying a more
187 robust solution.

188

189 We downloaded all GBIF observations of species with less than 5% of their range
190 covered by the KBA and PA networks ($n = 879$). We only used observations that
191 were made after 1990, because they have in general higher quality and are more
192 informative about the present occurrence locations of the species (Ficetola et al.,
193 2014). Because using a simple point occurrence would have given unrealistic weight
194 for the exact position where the species were observed, we made 25 km buffers
195 around occurrence points of each species to approximate species movement and
196 data uncertainty. This buffer size is within typical dispersal distances for terrestrial
197 vertebrates (Saura et al., 2018). For simplicity we decided to use only single
198 buffering distance for all species, because the aim of the buffers was more to simply
199 account for uncertainty in the location of the species and not to realistically model

200 species dispersal. We intersected the buffered occurrence rasters with species
201 ranges to remove observations outside natural ranges. We used these layers as an
202 additional input in the priority analysis of areas outside the KBA network to
203 upweight areas with confirmed species presences (Moilanen et al. 2006). There
204 were 7,555 observations of 104 species (775 of the 879 species were lacking
205 observations in GBIF) (GBIF, 2017) (Table A1, Fig. A1).

206

207 **2.5. Other data**

208 To estimate pressures from human activity to species within KBAs, we calculated
209 mean Human Influence Index (HII, WCS and CIESIN 2005) for all unprotected sites.
210 Human influence index has been shown to correlate with multiple factors relevant
211 for conservation (Hand et al. 2014, Safi & Pettorelli 2010, Yackulic et al. 2011).
212 Furthermore, we used World Bank's country income classifications for the 2019
213 fiscal year (World Bank, 2019) to explore how responsibility for the priority areas is
214 divided between countries having different resources for conservation.

215 .

216 **2.6. Prioritization analyses**

217 We used the Zonation v4 conservation prioritization software (Di Minin et al., 2014,
218 Moilanen et al. 2014), to produce a global priority ranking of the unprotected KBAs
219 for expanding the global PA network and to identify priority areas to work as
220 candidate sites for expansion of PA network outside KBAs The analyses were run

221 using the additive benefit function method, which aims to minimize the aggregate
222 extinction risk over all species (Moilanen, 2007). Robustness of the results with
223 respect to prioritization method was analyzed and confirmed, and is discussed in
224 supplementary material (Fig A4).

225

226 Prioritization covered all terrestrial areas in a hierarchical analysis, in which
227 highest, medium, and lowest priorities were forced into the current PA network,
228 unprotected KBAs, and the rest of the landscape, respectively (see e.g. Pouzols et al.
229 2014 for hierarchic analysis). In this structure, the highest priority unprotected
230 KBAs complement the current PA network area-efficiently in terms of balanced
231 coverage of threatened amphibian, bird and mammal ranges. To identify priority
232 areas for PA expansion outside the KBA network we focused on the third level of the
233 analysis hierarchy. These areas are considered in the analysis only after species
234 representation within the PA and KBA networks has already been accounted for and
235 thus highest priority is given to areas that best complement the species composition
236 of PA and KBA networks. In this third hierarchy level, each grid cell of the rasterized
237 data was used as an individual planning unit. Zonation produces a continuous
238 priority ranking of the full landscape (including unprotected KBAs and grid cells
239 outside of them). In this article we focus on the highest ranking cells that would
240 together with the priority KBA bring the PA coverage up to 17%, which is set as a
241 target for PA coverage in the Aichi targets (UNEP-CBD, 2010). Finally, to compare

242 the performance of the KBA based-solution to unrestricted expansion of the PA
243 network, we did a two-level, grid-based, hierarchical analysis expanding from the
244 present PA network to the rest of the landscape. Analysis variants are summarized
245 in Table 1.

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246 **Table 1:** Zonation analysis variants. PAN: protected area network, PA: protected area,
 247 KBA: Key biodiversity area.

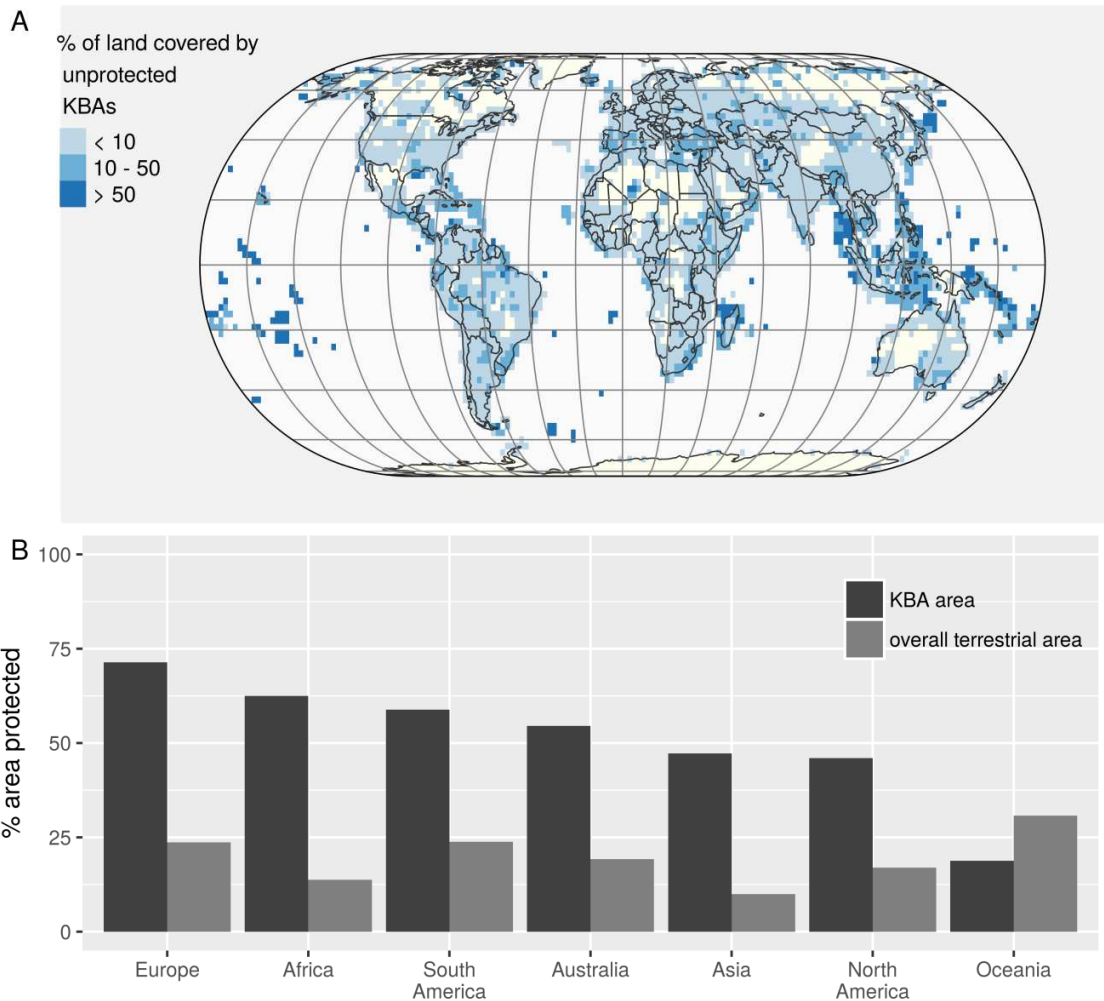
Analysis number and name	Purpose	Data used in prioritization (planning units)	Ranking hierarchy
1) Free PAN expansion	Create a grid-cell-based PA expansion ranking to help evaluate the performance of the KBA restricted solution	Threatened species (1' grid cells)	1. PAs (highest), 2. rest of the landscape
2) Priority KBAs for PAN expansion	Rank unprotected KBAs in terms of importance for improving the representation of the study species in PAN	Threatened amphibians, birds and mammals (unprotected KBAs)	1. PAs, 2. KBAs, 3. rest of the landscape
3) Priority areas for PA expansion outside the KBAs	Identify priority areas (1' grid cells outside PAN and KBAs) for potential expansion of the KBA and PA networks	Threatened species, data deficient species, and GBIF observations of gap species (1' grid cells)	1. PAs, 2. KBAs, 3. rest of the landscape

248

249 **3. Results**

250 **3.1. KBA and protected area coverage**

251 KBAs covered 8.85% (~12 000 000 km²) of the world's terrestrial surface. Overall,
252 55.8% (~6 700 000 km²) of KBA surface area was protected, leaving 10,430 KBAs
253 with at least some unprotected parts. Unprotected parts of the KBA network
254 (planning units of the analysis 2 of table 1), covered approximately 3.6% (5 300 000
255 km²) of the world's terrestrial surface and were spread fairly evenly, albeit with
256 some larger gaps especially in large desert areas and the Arctic (Fig. 1A). On all
257 continents except Oceania, the KBAs were better covered by protection than the
258 landscape on average (Fig. 1B).



259

260 **Fig. 1:** Distribution of unprotected terrestrial KBAs (A) and protected area coverage
 261 of KBAs by continent (B). Panel A shows the spatial distribution of unprotected
 262 terrestrial KBAs. Darker colors indicate higher coverage of the terrestrial surface by
 263 unprotected KBAs. Terrestrial areas without any unprotected KBAs appear white.
 264 These areas cannot be reached by prioritization analysis that is limited to
 265 unprotected KBAs only. In panel B the bars show the percentage of the total KBA
 266 area that is covered by protected areas (dark gray) and overall protected area
 267 coverage of the terrestrial areas (light gray). In all continents but Oceania protected
 268 area coverage of KBAs is higher than overall protection level of terrestrial areas.

269

270 PAs covered 14.3% of the terrestrial surface with a mean coverage of 32.5% of the
271 ranges of the study species; 302 (6.29%) species had their ranges fully protected
272 (Table 2). However, 672 (14%) species completely lacked coverage in the PA
273 network and these are hereafter called gap species. Together, PAs and KBAs covered
274 17.9% of the terrestrial surface and, on average, 54.3% of threatened species
275 ranges, leaving only 124 (2.58%) gap species and having full coverage of 828
276 species ranges. As Data Deficient species are not used in the identification of KBAs, it
277 is not surprising that the ranges of Data Deficient species are not well covered by
278 KBAs (Table 2). In fact, 19.9% of Data Deficient species had their whole (by
279 definition poorly known) range outside the PA and KBA networks.

280 **Table 2:** Coverage of the species ranges by different areas of interest. The number
 281 of analysis variant corresponds to the numbering in Table 1. Mean refers to mean
 282 percentage of species ranges covered. Gap species and full species refer to number
 283 of species completely missed by the solution and number of species ranges fully
 284 covered by the solution, respectively (percentage of all species in the group in
 285 parentheses). The values for "PAN and all KBAs" give the maximum species coverage
 286 that is reachable within the protected area and KBA networks. The size of the
 287 outside KBAs priority area is 2.1% of the terrestrial surface, which would together
 288 with top 10 % of unprotected KBAs increase the global PA coverage to 17%. PAN
 289 refers to protected area network and KBA to Key Biodiversity Areas.

Area of interest (number of analysis variant used)	Threatened species			Data Deficient species		
	mean	gap spp (%)	full spp (%)	mean	gap spp (%)	full spp (%)
PAN (1)	32.5	672 (14)	302 (6.29)	30.3	681 (29.28)	238 (10.23)
PAN and top 10 % of unprotected KBAs (2)	50.3	162 (3.38)	775 (16.15)	36.7	562 (24.16)	320 (13.76)
PAN and all KBAs (2)	54.3	124 (2.58)	828 (17.25)	42.7	463 (19.91)	372 (15.99)
PAN and unrestricted expansion areas (1) (same size with top 10 % of unprotected KBAs)	62.3	26 (0.54)	1324 (27.58)	38.2	515 (22.14)	332 (14.27)
PAN, KBAs and priority areas outside KBAs (3)	82.3	2 (0.04)	2445 (50.95)	86.6	0 (0)	1484 (63.8)

290

291

3.2. Priority areas for PA expansion within unprotected KBAs

292 We found that the highest ranking unprotected KBAs would be very effective in
293 improving the representation of species ranges within the PA network (Fig. A3).
294 Zonation produces a continuous ranking of the priority areas, but from here on, we
295 focus on 10% of the highest ranking unprotected KBAs, which are referred to as top
296 priority KBAs. The arbitrary 10% cut off value was chosen for communication
297 purposes, but it also falls to the period in the priority ranking where the benefit of
298 including new areas to the priority set, measured as species ranges covered,
299 decreases quickly (Fig A3). The exact priority ranking of all unprotected KBAs is
300 provided in a supplementary file (supplementary file B). The results of the analysis
301 (2) identifying the priority KBAs for PA expansion show that by protecting the top
302 priority KBAs (0.36% of terrestrial area) it is possible to increase the mean coverage
303 of ranges of threatened species by 17.8 percentage points while decreasing the
304 number of gap species from 672 (14.00%) to 162 (3.38%) and increasing the
305 number of species fully covered from 302 (6.29%) to 775 (16.15%). As expected,
306 the mean coverage of threatened species ranges was higher for the unconstrained
307 solution than for the one limited to using KBAs as expansions (Table 2).
308
309 Priority KBAs for PA network expansion consisted of 1,882 individual sites, with
310 median size of 76.2 km², which is close to the average size of the unprotected KBA
311 sites (median 82.9 km², Kruskal-Wallis test, df = 1, p = 0.09) (See Fig A5 for a
312 breakdown of the size distribution) and considerably larger than average size of
313 protected areas (median 0.5 km², Kruskal-Wallis test, df = 1, p < 0.001). Most

314 priority KBAs were located at lower latitudes (Fig. 2), especially in Central America,
 315 the Amazonian Andes, Eastern Madagascar, and Southeast Asia (Fig. 2). Priority
 316 KBAs had, on average, more establishment criteria attached to them (2.49 criteria /
 317 area) compared to the other unprotected KBAs (1.84 criteria / area). Criteria
 318 focusing on species rarity were more common in the priority sites than among the
 319 other unprotected KBAs (Table 3).

320

321 **Table 3:** *Percentage of KBAs with different establishment criteria. Different groups*
 322 *refer to priority KBAs, all unprotected KBAs and all KBAs. The sum of the proportions*
 323 *within the groups is higher than one because single KBA can be established based on*
 324 *multiple different criteria. The criteria refer to the reasons why the KBA was*
 325 *established for as mentioned in the KBA dataset (KBA Partnership, 2016): CR/EN =*
 326 *Important for Critically endangered or Endangered species, VU = Important for*
 327 *Vulnerable species, endemic = Important for endemic species, migratory birds or*
 328 *congregations = areas important for bird migration or other species seasonal*
 329 *congregations, other = mixture of all other criteria (see for example IUCN 2016a for*
 330 *full description).*

criteria	priority	unprotected	all
CR/EN	0.71	0.38	0.34
VU	0.63	0.43	0.40
endemic	0.73	0.31	0.28
migratory birds –	0.15	0.33	0.33

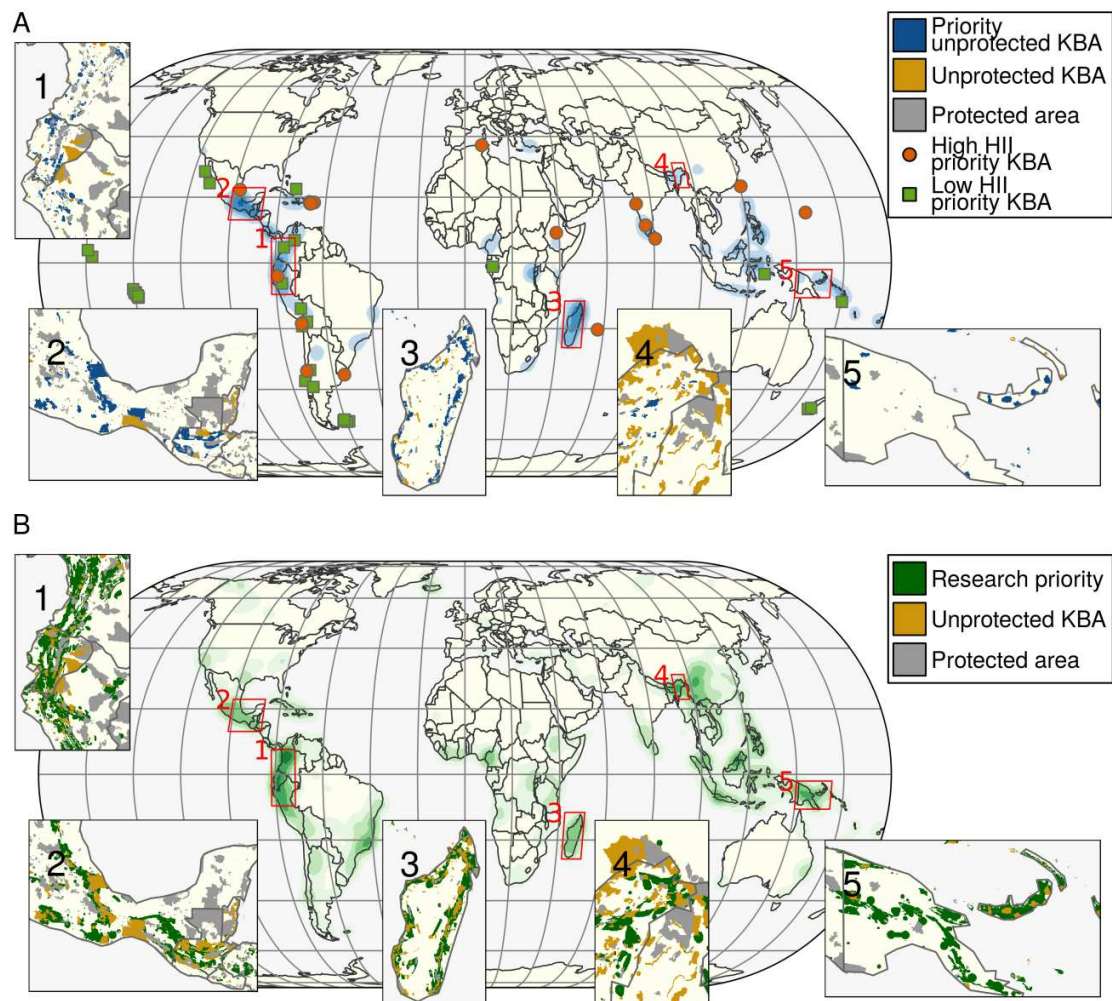
congregations

other	0.26	0.39	0.42
-------	------	------	------

331

332 Median HII within the priority KBAs was 18.20 (theoretical maximum 72), which is
 333 slightly lower than the median of all unprotected KBAs (19.30, Kruskal-Wallis test,
 334 $df = 1, p < 0.001$). Specifically, 25 priority KBAs without any protection had mean
 335 HII value ranking among the highest 5% of all unprotected KBAs ($HII > 40.20$) and
 336 27 among the lowest 5% ($HII < 5$). Figure 2A shows priority KBAs with the highest
 337 HII as red dots and those with the lowest HII in as green dots (See tables A3 and A4
 338 for more information about these KBAs).

339



340

341 **Fig. 2:** Priority KBAs and priority areas outside KBAs identified in the analysis. The
 342 main map in panel A highlights the areas that contain a high density of priority Key
 343 Biodiversity Areas (KBAs) for protected area expansion based on threatened species
 344 ranges. Panel A also shows priority KBAs with high (> 40.20) and low (< 5) human
 345 influence index (HII) as red and green dots respectively. Panel B highlights locations
 346 with a high density of outside KBA network priority areas that are potential areas
 347 for effective protected area and KBA network expansion. Inset maps show detailed
 348 arrangements of priority KBAs, priority areas outside KBAs, unprotected KBAs and
 349 protected areas in selected locations. A full resolution map of priority KBAs is
 350 provided in the supplementary material (supplementary file C: analysis outputs).

351

352 **3.3. Priority areas outside the KBA network**

353 The distribution of priority areas outside the KBA network was similar to the
354 pattern of KBA priorities (Fig. 2): the Amazonian Andes, the Atlantic coastal forest in
355 Brazil, Western Africa, continental Southeast Asia and Papua New Guinea were
356 highlighted as areas with highest potential for KBA and PA network expansion. If
357 placed under conservation management, these areas would increase the
358 coverage of threatened species ranges by an additional 28 percentage points
359 compared to PAs and KBAs alone (from 54.3% to 82.3%) and leave only 2 species
360 without any conservation coverage (Table 2). The priority areas outside the KBA
361 network also overlap with ranges of all Data Deficient species with a very high mean
362 coverage of 86.6%. Accounting for Data Deficient species and GBIF observations in
363 the analysis did not alter the general global priority pattern (Fig. A2). Its effect on
364 representation of threatened species was also negligible, but representation of Data
365 Deficient species ranges and GBIF observations were increased drastically (Table
366 A2).

367

368 Table 4 shows that compared to other species groups amphibians might benefit
369 relatively more from grid based prioritization that allows selecting sites also outside
370 the unprotected KBAs (analysis 1). This can be noted from a relatively larger
371 increase in the representation of amphibian species in the grid based PA expansion

372 as compared to the analysis that is restricted to unprotected KBAs (analysis 2). On
 373 average, amphibian ranges are covered better than other species groups in both
 374 analyses and in protected areas. This is probably caused by relatively small range
 375 sizes of amphibians (median 39600 km²) which are easier to cover in the
 376 prioritization analyses compared to larger ranges of mammals and birds (median
 377 675 000 km² 855 000 km² respectively).

378 **Table 4:** Mean percentage of species ranges covered by protected areas and priority
 379 sets identified by grid-based (analysis 1 in table 1) and KBA-based analyses
 380 (analysis 2 in table 1). Priority sets refer to 10% highest ranking KBAs and similar
 381 area of highest ranking grid cells. Increase in the mean representation of species
 382 ranges in the free approach compared to the KBA based approach is calculated as:
 383 free priority / priority KBAs x 100.

Species group	Coverage of ranges (mean %)			Increase in representation of the free approach compared to the KBA based approach (%)
	PAs	priority KBAs	free priority	
Amphibians	35	23	42	186
Birds	29	15	22	145
Mammals	30	13	20	151

384

385 **3.4. Country responsibility of the priority areas**

386 Priority unprotected KBAs were found in 118 (47%) countries. The priority KBAs
 387 were highly concentrated, so that six countries, Ecuador, Indonesia, Madagascar,
 388 Mexico, Peru and Philippines, alone covered 45.2 % of all priority KBA surface. In
 389 addition, 73.3% of the priority KBA surface areas was located in middle income

390 countries (n = 102). Low income countries (n = 34) covered 19.4%, while high
391 income countries (n = 79) covered 7.1% of the total surface area of priority KBAs.
392 High correlation between the proportion of priority KBAs with proportion of
393 priorities outside KBAs (identified with analysis 3, Pearson $r = 0.79$, $p < 0.001$) and
394 unrestricted priorities (identified with analysis 1, Pearson $r = 0.93$, $p < 0.001$),
395 suggest that same countries bear high responsibility of the overall global
396 conservation priorities despite the identification method. Full data is provided as
397 supplementary material (Table A5). The fact that representation of data deficient
398 species and GBIF observations is improved without changes in the global priority
399 pattern suggests that the improvement is caused by local level shift in the priority
400 pattern.

401

402 **4. Discussion**

403 **4.1. Unprotected KBA priorities for global protected area** 404 **network expansion**

405 Recent analyses of global conservation priorities have reported high potential for
406 increasing coverage of species ranges with small additions to the PA network
407 (Butchart et al., 2015, Pouzols et al., 2014, Venter et al., 2014). These analyses have
408 successfully directed attention towards areas where protection would benefit global
409 biodiversity the most. Nevertheless, as the analyses have been mostly based on large
410 grid cells or ecoregions as planning units, the priority areas could not necessarily be

411 protected as such. Deciding about which particular areas to protect within the
412 identified priority areas would require additional knowledge about local
413 circumstances. Selecting areas for protection within the priority sites could lead to
414 unexpected outcomes because the areas that are important for biodiversity might
415 not be available for conservation purposes or the priority grid cells might not alone
416 be suitable for sustaining important ecological processes (Hurlbert & Jetz, 2007).

417

418 Our results show that large gains (18 percentage points) in species representation
419 could also be achieved with very limited area if the PA network is expanded to
420 priority KBAs (0.36% of land area, Table 2), which clearly are units fit for
421 conservation. To meet the 17% area coverage target, three-quarters of the
422 unprotected KBAs should be protected. This would further increase conservation
423 coverage of threatened species, although not as effectively as selection that
424 combines unprotected KBAs and freely selected areas (Table 2). It is good to
425 remember that at the same time as countries are pursuing to reach the PA coverage
426 targets set by the CBD, many PAs are lacking funds for proper management
427 (Waldron et al., 2017). To truly slow down the biodiversity crisis, resources should
428 also be targeted to improving the management of existing PA network (Waldron et
429 al., 2017).

430

431 Using global conservation priorities to inform local conservation action is
432 challenging. However, using unprotected KBAs as planning units, as we did here, can
433 help reduce the gap between global priorities and local level conservation actions.
434 We provide a list of priority KBAs as a supplementary table (see supplementary
435 table A3 for the top priority KBAs under high pressure and supplementary file B for
436 full list of unprotected KBAs) and hope that it could be used to draw attention to
437 those unprotected KBAs that are the most valuable from the point of view of global
438 conservation priorities. Our results also demonstrate that it is possible to do this
439 without compromising overall representation of biodiversity.

440

441 The global pattern of priority KBAs agrees with priorities identified in previous
442 global analyses (Butchart et al., 2015; Pouzols et al., 2014; Venter et al., 2014) and
443 with the priority pattern outside KBAs. This is caused by the underlying richness
444 pattern of restricted range vertebrates (Jenkins et al., 2013), which strongly drives
445 all global conservation prioritization analyses that account for species ranges. The
446 fact that most of the global priority areas for conservation are situated in the global
447 south where funds for conservation might be scarce (middle or low income
448 countries), sets additional challenge for moving from plans to implementation.
449 Therefore mobilization of resources at the global level, which is the focus of Aichi
450 target 20 (UNEP-CBD, 2010), should also have high importance for post 2020
451 conservation plans.

452

453 Nevertheless, even within regions with the highest density of global priority KBAs
454 and priority areas outside KBAs, there were individual KBAs that were not included
455 in the priority KBA set, indicating variation in the importance of KBAs at the
456 regional level (Fig. 2). The areas that are ranked high compared to the surrounding
457 areas are likely to contain small ranged species that cannot be covered anywhere
458 else, whereas areas ranking lower than their surrounding areas might only contain
459 species that are already well covered by PAs or other KBAs.

460

461 **4.2. Priority areas outside KBAs needed to cover all species**

462 The aim of the grid-based priority analysis of areas outside the KBA network was to
463 focus attention to sites that might well be valuable for improving the
464 complementarity of the global PA and KBA networks, but that simply cannot be
465 reached with an analysis that is restricted to KBAs only. Compared to the
466 unprotected KBAs, which are delineated for conservation, these sites cannot be
467 assumed to be suitable for conservation as such. The availability of these sites for
468 conservation and possible delineation of new PAs should be based on ground
469 surveys, with site level information about actual species occurrences, habitat
470 quality, costs and social factors. (Margules & Pressey, 2000).

471

472 Typically countries with many priority KBAs had also many priority areas outside
473 KBAs. These countries have high species diversity and many species with very
474 restricted ranges. Of all countries, Madagascar had the largest share of the global
475 priority KBA surface area within its borders (12%) (Table A5). It also covered 4% of
476 the total global surface area of the priority areas outside the KBAs (the eight highest
477 of all countries). Indonesia had the largest share of surface area of global priority
478 areas outside KBAs (10%) and third highest share of global priority KBA surface
479 area (8%) (Table A5). Some countries like Fiji had most of its surface area assigned
480 to either priority KBAs or priority area outside KBAs, but responsibility of the
481 overall global priority remained small due to the small size of the country.

482

483 One of the most notable differences between the prioritization approaches was the
484 low density of top priority KBAs combined with high priorities outside KBAs in
485 mainland Papua New Guinea, which is commonly recognized as a global
486 conservation hotspot (Jenkins et al., 2013). This area has very few PAs and few
487 unprotected KBAs available for selection, increasing its importance as an area for
488 establishing new PAs or KBAs. On the other hand, some large countries like China,
489 had considerably high responsibility of the priority areas outside KBAs, but at the
490 same time, although unprotected KBAs would be available, relatively few priority
491 KBAs (Table A5). This could indicate that there are many threatened species that
492 are missed by both PA and KBA network. In contrast, some countries, like

493 Guatemala, had considerably lower overall responsibility of the priority areas
494 outside KBA network than inside KBA network (Fig. 2), suggesting that
495 comparatively many species in that area are already well covered either by KBAs or
496 the PA network.

497

498 KBA standard suggest quantitative analyses as a one option to identify new KBA
499 sites (IUCN, 2016a). Therefore, the priority sites identified here could also be used
500 to indicate possible areas for KBA expansion. Nevertheless, it should be noted that
501 because our priority analyses were based on species data only, they cover only one
502 part of KBA criteria (IUCN, 2016a) and thus their usability for KBA expansion is
503 limited to identifying sites that are valuable for threatened species (KBA criteria A1,
504 IUCN, 2016a). Priority analyses that could account for other criteria such as,
505 migration and ecosystems would be needed to improve subsequent analyses for
506 KBA expansions.

507

508 The current KBA network is strongly based on the Important Bird Areas that have
509 also functioned as an inspiration for the whole KBA concept (Eken et al., 2004).
510 Although KBAs are currently identified with a broader biodiversity focus (IUCN,
511 2016a.), due to historical reasons, birds can still be expected to be better covered by
512 the network. This might lead to a situation where prioritization that is based on the
513 KBAs might favor bird species. In our analysis there are no strong signs of this, as is

514 shown by the relative large cover of other species groups by the priority KBAs and
515 KBAs as a whole (table 4). On the other hand, compared to birds and mammals,
516 amphibians would benefit more by not restricting the prioritization to KBAs. This is
517 because a large number of small ranged amphibians that are not covered by the KBA
518 network. Although the differences are not large, one should be cautious when using
519 KBAs for conservation prioritization of species groups that have not been at the
520 focus of the KBA identification work.

521

522 Priority areas outside the KBA network are especially important for Data Deficient
523 species, many of which are missed by priority KBAs (Table 2). This is not surprising,
524 because the KBA standard particularly emphasizes the importance of confirmed
525 knowledge about species occurrences (IUCN, 2016a). Because Data Deficient species
526 might well be rare and have restricted ranges (Bland et al., 2015; Trindade-Filho et
527 al., 2012) and because it was possible to account for them without compromising
528 representation of endangered species (Table A.2), including them in analyses for
529 expanding PAs outside the KBA network is the safest bet.

530

531 We also found that GBIF species observations can be accounted for in the priority
532 analyses without notable loss in coverage of species ranges (Table A.2). Because
533 there were only few observations per species and because GBIF observations are
534 known to be taxonomically and spatially biased (Meyer et al., 2015), global priority

535 setting cannot rely solely on them. For the same reasons, building reliable species
536 distribution models at global scale might be challenging (van Proosdij, 2015).
537 Nevertheless, we believe that species observations can safely be used to identify
538 areas with higher confidence of species presence in a analyses that are based on
539 species ranges only.

540

541 **4.3. Global analyses are restricted by data availability**

542 Our analysis aimed to efficiently increase the coverage of threatened terrestrial
543 vertebrate species. However it did not account for other ecological factors
544 influencing the KBA status of an area (IUCN, 2016a). This is reflected strongly in the
545 higher proportion of species occurrence-based establishment criteria within the
546 present priority KBAs. This observation is consistent with Di Marco et al. (2016),
547 who found that higher ecological irreplaceability of KBAs was associated with
548 presence and number of restricted-range species. Other establishment criteria such
549 as importance for species migration or importance for species that were not
550 accounted for in this analysis can partly explain why some KBA seems to contribute
551 only little to the global conservation coverage of species ranges in the present
552 analysis. Further, our analyses give highest value for sites that are important for
553 many species at the same time. In this type of approach, areas that might be
554 critically important for some individual species, but are otherwise species poor,
555 might not appear as high priority. It is important to note that these areas might still

556 be valuable for conservation and need protection, although they are not the areas
557 that are the most effective in enhancing the representativeness of the global
558 protected area network.

559

560 As a data-driven process, the outputs of conservation prioritization analyses should
561 be interpreted according to understanding about underlying data and methods.
562 Firstly, results only apply to taxa included in analysis, in this case threatened and
563 data deficient terrestrial birds, mammals and amphibians. The effectiveness of the
564 priority areas in covering other taxa should be treated cautiously since several
565 meta-analyses have reported low performance of between-taxa surrogates (de
566 Morais et al., 2018, Westgate et al., 2014). Thus, species groups with limited
567 distribution information at global level, such as invertebrates, could be given special
568 attention when new KBAs or PAs are established based on locally available data. On
569 the other hand, Surrogates are also likely to work better in prioritization studies
570 with broad extent (Lamoreux et al., 2006, Westgate et al. 2014) and large number of
571 species (Kujala et al., 2018) like this one. Nevertheless even at the global scale,
572 adding completely new taxa with many species is likely to cause some shifts in the
573 locations of the priority areas (Roll et al., 2017). Secondly, although being best
574 available, the range maps are known to have limitations as biological data (Di Marco
575 et al., 2017), although with large data sets effects of problems with individual layers
576 are strongly reduced (Kujala et al., 2018). It has also been shown that although the

577 importance of site for individual species might be difficult to determine from the
578 range maps the importance for biodiversity in general can be inferred more robustly
579 (Maréchaux et al., 2016). Our attempts to use GBIF data to improve the
580 prioritization also works towards improving the analyses based on the range maps,
581 but as there are only few observations available it is nowhere near solving the
582 problem. Thirdly, other factors like cost and threats can have large effect to the
583 priority pattern (Carwardine et al., 2008). We decided not to account for costs or
584 threats directly in our analyses, because we wanted to follow the approach taken by
585 the KBA standard and focus purely on identifying sites that are important for
586 species persistence at the global level (IUCN, 2016a). Therefore, although our
587 prioritization analyses are effective in terms of area and vertebrate species
588 representation, it might neither be the cheapest solution nor necessarily identify
589 areas having highest urgency for protection of other higher taxa.

590

591 Instead of inputting threat data directly to the prioritization analysis itself, we
592 highlighted the biodiversity priority sites with highest levels of human influence
593 within their borders as areas that might require action most urgently (Di Minin et al.
594 2019). KBAs with both high and low HII were evenly represented in the set of
595 priority sites making it possible to focus on sites with high or low pressure (Fig. A6).
596 KBAs with high human influence might need protection most urgently, but at the
597 same time these sites might be more expensive to protect because they area also

598 important for human activities. As a comparison, Venter et al 2014 used agricultural
599 opportunity cost and Butchart et al (2015) human population density as a cost layer
600 in their prioritization analyses. These factors are likely to be correlated with HII and
601 therefore drive the priority areas away from the areas with highest pressures. Our
602 view is that, because especially at the global scale, it might be difficult to say
603 whether the areas with high pressure and high cost should be preferred or avoided,
604 it is safest to simply identify areas with highest value for biodiversity and let the end
605 users to decide which sites to prefer.

606

607 Our results reflect the priority rank of the sites especially at the time of the analysis.
608 When new KBAs and PAs are established or new biological data becomes available
609 the priority pattern will be affected. For example if protection of some species is
610 improved in one location, the priority ranking of areas containing that species
611 elsewhere is likely to be reduced. Therefore, although the overall global priority
612 pattern is rather robust against small changes, if the ranking is used for local level
613 decision making, the analysis should be update whenever there are large changes to
614 the input data.

615

616 **4.4. Using KBAs improves the robustness of the prioritization**

617 Using KBAs as selection units sets additional constraints to the prioritization
618 analyses (Moilanen et al., 2009), which inevitably reduces their theoretical

619 performance compared to a grid-based analysis (Table 2). The reason for this is that
620 even the high priority KBAs will include some lower priority areas, and some areas
621 with high value are located outside the KBA network. Despite the lower theoretical
622 performance of the KBA based solution, it might be more effective for guiding the
623 expansion of the global PA network in practice. This is because the priority areas
624 identified in the pixel-based solutions are likely to contain areas that are not
625 suitable for protection due to factors that were not accounted for in the analysis. For
626 example information on current land use, land ownership (Di Minin et al., 2017) and
627 local ecological processes (Pressey et al., 2007) are important for establishing
628 protected areas, but cannot be accounted for directly in global scale conservation
629 prioritization analyses. Further, the pixel-based priority areas might be too small
630 and have shapes that are not suitable for effective management of protected areas.

631

632 On the other hand, based on the IUCN standards, the KBAs, and thus also the priority
633 KBAs, should be delineated in a way that allows effective management and
634 important ecological processes to be sustained within the areas (IUCN, 2016a).
635 Detailed local level knowledge about factors influencing the suitability of areas for
636 protection can also be accounted for already when the KBAs are delineated (IUCN,
637 2016a), making the priority KBAs more likely to be directly suitable for protection.
638 KBAs are shown to represent local biodiversity better than expected compared to
639 the surrounding landscape (Di Marco et al., 2016). This is likely to further improve

640 the actual performance of the KBA based solution by reducing commission errors
641 that are inherent for the global scale species data (Di Marco et al., 2017). Finally, in
642 addition to conservation value identified in this analysis, the priority KBAs are
643 already identified to be crucial for global biodiversity by the KBA method making
644 them a double priority and thus prime candidates for protection.

645

646 Nevertheless, it is important to notice that some species are totally missed by the
647 current KBA and PA networks (Table 2). To improve the coverage of these species,
648 the priority KBA approach should be supplemented with new KBAs and PAs that are
649 established in the priority areas outside of the KBA network (fig 2). Compared to
650 the areas identified in the priority KBA analysis, more field studies are needed in
651 these areas to confirm the presence of the species and to collect information about
652 local circumstances before any on-the-ground action can be taken.

653

654 **5. Conclusions**

655 Our objective was to rank KBAs in terms of how well they would complement the
656 current PA network and thus work as an effective expansion for it. Every KBA is by
657 definition important for biodiversity, and continues to be so, whether or not it was
658 included in the priority sets identified here. We are not suggesting that protection is
659 the only solution for safeguarding the conservation value of KBAs. In many cases,
660 the biodiversity value of a KBA could well be maintained with other conservation

661 actions (IUCN, 2016a). Nevertheless, most of the top priority KBAs and priority
662 areas outside KBAs are located in regions where pressures on biodiversity are
663 expected to intensify (Tilman et al., 2017). Therefore, strengthening the
664 conservation status of these areas that overlap with the ranges of many globally
665 threatened species might be a worthwhile investment for the future.

666

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671

672 **7. Data and data manipulation**

673 The code for data manipulation, analysis and production of this document is
674 provided in
675 <https://github.com/PKullberg/Priorities for KBA research and protection>*[the link*
676 *to the final version will be made public after publication]*. Zonation software is
677 available from [www.helsinki.fi/en/researchgroups/digital-geography-](http://www.helsinki.fi/en/researchgroups/digital-geography-lab/software-developed-in-cbig)
678 [lab/software-developed-in-cbig](http://www.helsinki.fi/en/researchgroups/digital-geography-lab/software-developed-in-cbig). All relevant data is derived from published sources
679 indicated by the citations in the methods section.

680 **8. Supplementary files**

681 Supplementary file A: additional figures and tables

682 (https://drive.google.com/file/d/17gCMWhXQC2uMaINLH2j_1Bd0-

683 [nr0SqCQ/view?usp=sharing](https://drive.google.com/file/d/17gCMWhXQC2uMaINLH2j_1Bd0-nr0SqCQ/view?usp=sharing)).

684 Supplementary file B: full list of KBAs, available at

685 https://drive.google.com/file/d/1wd7dbb8LkbCG1V1wqHWz4wd6513i5_yR/view?

686 [usp=sharing](https://drive.google.com/file/d/1wd7dbb8LkbCG1V1wqHWz4wd6513i5_yR/view?usp=sharing) (*This will be moved to a permanent repository after acceptance*).

687 Supplementary file C: The result files of the priority analyses including the detailed

688 priority maps can be downloaded from:

689 https://drive.google.com/file/d/1j17jb_rh3EEFSYim7MF1IIRfbeiBjrLR/view?usp=s

690 [haring](https://drive.google.com/file/d/1j17jb_rh3EEFSYim7MF1IIRfbeiBjrLR/view?usp=sharing). (*This will be moved to a permanent repository after acceptance*)

691

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Highlights

1. We prioritize unprotected KBAs for effective expansion of the global PA network
2. Analysis is based on terrestrial vertebrate ranges and uses KBAs as planning units
3. Priority KBAs covered biodiversity broadly but some species were missed
4. Restricting expansion only to KBAs lowers the representation of biodiversity
5. Priority KBAs with few critical additions are a good complement for the PA network

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