

Coevolution and coexistence

1 **Predator co-evolution and prey trait variability determine species coexistence**

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26 **Abstract**

27 Predation is one of the key ecological mechanisms allowing species coexistence and influencing
28 biological diversity. However, ecological processes are subject to contemporary evolutionary change,
29 and the degree to which predation affects diversity ultimately depends on the interplay between
30 evolution and ecology. Furthermore, ecological interactions that influence species co-existence can be
31 altered by reciprocal co-evolution especially in the case of antagonistic interactions such as predation
32 or parasitism. Here we used an experimental evolution approach to test for the role of initial trait
33 variation in the prey population and co-evolutionary history of the predator in the ecological dynamics
34 of a two-species bacterial community predated by a ciliate. We found that initial trait variation both at
35 the bacterial and ciliate level enhanced species coexistence, and that subsequent trait evolutionary
36 trajectories depended on the initial genetic diversity present in the population. Our findings provide
37 further support to the notion that the ecology-centric view of diversity maintenance must be
38 reinvestigated in light of recent findings in the field of eco-evolutionary dynamics.

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48 1. Introduction

49 Natural selection acts on the fitness variance of phenotypes [1–3], and adaptive evolution is predicted
50 to be enhanced in populations with high genetic diversity [4]. Adaptive evolution can lead to niche
51 divergence and thus facilitate coexistence in competitor or predator-prey communities [5–7]. Classic
52 competition theory predicts that species coexistence is possible when intraspecific competition is
53 stronger than interspecific competition [5]. However, owing to contemporary evolution, the impact of
54 competitors on each other might not be constant [8]. For example, the high number of coexisting species
55 in microbial communities [9] in even simple environments [10–13] might be explained by high levels
56 of within-species clonal diversity [5,14,15] and result in rapid evolution to use underexploited or new
57 ecological niches [13]. Recent work has shown that traits and species interactions in microbial food
58 webs can be altered by evolutionary change, as *de novo* mutations and changes over time in the
59 frequencies of genotypes from standing genetic variation can occur at the same temporal scale as
60 ecological processes [16–18]. It has been previously shown that consumers such as predators or
61 parasites can have significant indirect effects on the outcome of competition and thus on the
62 maintenance of species diversity [19] by compensating for differences in traits [5,20,21]. Direct
63 interaction between competing species might also change due to rapid evolution, resulting in resource
64 use divergence [13].

65 In microbial communities, competitive interactions are common [22–25], which can lead to rapid
66 exclusion of community members under standard conditions [21]. Recent studies have proposed that
67 coexistence and escape from competitive exclusion is facilitated by evolutionary change in between-
68 species interactions [13,26–28], resulting in e.g. cross-feeding [26] or a niche shift to underexploited
69 resources [29–31]. In most bacterial experiments, monoclonal isolates are assembled [13,21,23,26,32]
70 and any evolutionary change is based on *de novo* mutations [33–35], potentially imposing constraints
71 on evolution [4]. Our aim was to study the role of initial genetic variation on co-existence as the
72 associated phenotypic variation might affect competition [36]. Our study system consisted of two
73 bacterial species, *Escherichia coli* and *Pseudomonas fluorescens*, competing for shared resources and
74 consumed by a keystone predator, the ciliate *Tetrahymena thermophila*. To investigate if genetic
75 diversity can promote coexistence, we compared community dynamics with monoclonal *P. fluorescens*

76 populations to genetically diverse populations obtained by pooling different *P. fluorescens* clones. The
77 genetic diversity is represented by differences in phenotypic traits, including growth capacity and level
78 of defence against ciliate predation. To investigate if diversity in the predator population also has an
79 effect, we added a population containing diverse ciliate phenotypes obtained by pooling ciliates that
80 had coevolved with either of the two bacterial species. The reason for the two predator treatments is
81 that we hypothesise that if predation is a key factor allowing our two bacterial species to co-exist,
82 evolutionary adaptation in the predator allowing for more efficient predation might further facilitate the
83 co-existence [37].

84 In our experiment, we tracked the community dynamics and the evolutionary outcome of both
85 bacterial species when i) growing without a predator, ii) with a naïve ciliate predator and iii) with
86 coevolved predators. These predation treatments are hereafter rereferred as 'no predation', 'naive' and
87 'coevolved' treatments correspondingly. We also manipulated the genetic diversity of the *P. fluorescens*
88 population in a full factorial design. We used isogenic lines of *P. fluorescens* (and *E. coli*) to inoculate
89 the experiments with minimum standing genetic variation as control (population hereafter: 'ancestor').
90 Further, we increased the genetic diversity of *P. fluorescens* by adding evolved diverse populations
91 from previous experiments ('full-diversity'), or artificially assembled a population consisting of a subset
92 of clones ('high-diversity'). We assessed the variability in interaction between the two bacterial species
93 by measuring competitive dynamics and the level of coexistence. Ecological dynamics in population
94 size were followed for 16 days and evolutionary dynamics were estimated by testing whether the
95 bacteria evolved defence against the ciliate (measured as prey defence level), as well as by estimating
96 bacterial fitness (measured as growth capacity).

97 We found that manipulating within-species bacterial diversity affected the frequencies of the two
98 competitors over time but this effect depended on the presence and evolutionary history of the ciliate.
99 Notably, the highest frequency of *P. fluorescens*, expected to be the inferior competitor, was observed
100 in communities with standing genetic variation in both the bacteria and the ciliates. Our results show
101 that the relative contribution of evolution (temporal changes in growth capacity and/or defence traits)
102 and ecology (competitive interactions and predation) to changes in the frequency of the two bacteria
103 over time was strongly dependent on standing genetic variation.

104

105 **2. Methods**106 **(a) Model system**

107 We constructed our communities using two bacterial species and a ciliate predator, adopting a
108 previously utilized model system [21]. The two bacterial species are *Escherichia coli* ATCC 11303 and
109 *Pseudomonas fluorescens* SBW25 cultured in 1% King's B (KB) liquid culture medium. In general, the
110 ancestral *E. coli* strain seems to be the dominant competitor in co-cultures but is more limited by
111 predation than *P. fluorescens* (electronic supplementary material, Fig. S1). As a generalist predator,
112 capable of consuming both bacterial species, we used the ciliated protozoan *Tetrahymena thermophila*
113 CCAP 1630/1U. Prior to the experiments, all bacterial stocks were kept at -80°C and ciliate stocks
114 were cultured axenically in proteose peptone yeast extract (PPY) medium containing 20 g of proteose
115 peptone and 2.5 g of yeast extract in 1 L of deionized water.

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117 **(b) Obtaining trait diversity**

118 For *P. fluorescens* and the ciliate predator, we manipulated genetic diversity by combining samples
119 isolated from a long-term predator selection experiment (LTPE). The LTPE was started using a single
120 colony *P. fluorescens* SBW25 and *E. coli* ATCC 11303, and an axenic culture of the ciliate *T.*
121 *thermophila* 1630/1U (CCAP). Material from only these two selection lines was used in the current
122 experiment. Each bacterial strain was cultured alone and together with the ciliate (three replicates each)
123 in 20 ml glass vials containing 6 ml of 5% KB medium, with 1% weekly transfer to fresh medium.
124 Cultures were kept at 28°C ($\pm 0.1^{\circ}\text{C}$) with shaking at 50 r.p.m. Every four transfers (28 days), bacterial
125 and predator densities were estimated using optical density as a proxy for bacterial biomass and direct
126 ciliate counts as described previously [38], and samples were freeze-stored with glycerol at -20°C for
127 later analysis. This experiment had been running for 20 months when we isolated the populations for
128 the current experiment.

129 We isolated coevolved ciliates from each of the six LTPE *P. fluorescens* and *E. coli* lines (three
130 replicate lines each) and pooled them together in equal densities to obtain a diverse ciliate population
131 referred as "coevolved ciliates". This mix of ciliates was cultivated axenically in PPY until the start of

132 the experiment. The full-diversity population of *P. fluorescens* was harvested by mixing all three
133 replicate populations from samples freed from live ciliates through freeze-storage at -20°C (ciliates do
134 not survive under these conditions). For ancestral and high-diversity populations, we isolated individual
135 colonies of the ancestral strains and from two time points in the LTPE, after six months and after 20
136 months, using Tryptone Bile X-Glucuronide agar (TBX, Sigma-Aldrich) and CFC agar plates (CFC
137 supplement: 10 mg of ceftrimide and fucidin and 50 mg cephalosporin in 1 l of PPY agar). We
138 determined the position of each clone in trait space (see below) comprising growth capacity and level
139 of anti-predatory defence. For both ancestral bacterial strains, we isolated 10 colonies and picked one
140 isolate representing the ancestral trait space mean. We initially isolated and characterised a pool of 200
141 clones and picked 20 clones among them representing a broad range of phenotypes in growth capacity
142 and defence against predation. The high-diversity population was established by randomly combining
143 10 out of these 20 clones.

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145 **(c) Determining position in trait space**

146 For individual clones from both bacterial species, we determined growth capacity and defence level.
147 For these measurements, we used the Bioscreen C spectrophotometer (Growth Curves AB Ltd,
148 Helsinki, Finland) to estimate the optical density of growing bacterial samples (100 wells) at 5 min
149 intervals for 48 hours. Frozen samples were revived in fresh medium and allowed to acclimatize for 24
150 hours, after which they were pin-replicated to fresh conditions (1% KB) in Bioscreen-compatible
151 honeycomb plates. These plates were incubated at 28°C under constant shaking in the Bioscreen device.
152 As a proxy for biomass yield, we calculated the area under the curve to obtain growth capacity for each
153 clone. After 48 hours, ciliates were added to the samples to estimate biomass loss due to predation.
154 Comparing change in the bacterial biomass of control treatments without ciliates with treatments
155 containing ciliates allowed us to measure the loss of bacterial biomass due to predation. Comparing
156 between individual clones allowed us to estimate which clones are well defended and which clones are
157 poorly defended. Briefly, ciliates were cultivated 5 days in advance in fresh PPY medium. The medium
158 was removed by centrifugation (2×8 minutes at 3300 r.p.m.) and the populations were starved
159 overnight in M9 salt solution. Initial *T. thermophila* cell densities were enumerated directly from live

160 subsamples (2.5 ml) using a compound microscope (Zeiss Axioskop 2 plus, Oberkochen, Germany)
161 and diluted to obtain 1000 cells ml⁻¹ inoculated to each microcosm. For control treatments, we filtered
162 ciliates out from the culture vial and added ciliate-free filtrate. The optical density of the samples was
163 tracked again for 48 hours, and the loss of biomass due to predation was estimated by comparing the
164 control with the predation treatment. This protocol allowed us to determine growth capacity and defence
165 level in the same population. For determining evolution, we picked 10 *E. coli* and 10 *P. fluorescens*
166 clones from each microcosm at the end of the experiment. For these clones, the evolved position in trait
167 space (i.e. growth capacity and defence level) was estimated using the protocol described above.

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169 **(d) Estimating ciliate growth on both bacterial strains**

170 To estimate ciliate performance, we isolated both the ciliates and the bacterial strains from the LTPE.
171 Briefly, we isolated the bacteria by freezer-storage which effectively killed all ciliates. Axenic ciliate
172 populations were obtained by culturing experimental populations to high density in PPY medium
173 containing 24, 50, 50 and 33 µg ml⁻¹ of the antibiotics kanamycin, rifampicin, streptomycin and
174 tetracycline, respectively. Axenicity was controlled for by plating on 50% PPY agar plates (on which
175 all experimental bacterial strains grow). Following this, ciliates were transferred to antibiotic-free PPY
176 medium and cultured to high density. For the growth assay, we grew both evolved bacterial strains in
177 5% KB to equal density based on optical density. Bacterial cells were centrifuged, and the medium was
178 replaced with M9 salts, preventing further bacterial growth. Ciliates from both coevolved lines were
179 grown in PPY to high density and the medium was replaced with M9 minimum medium. Ciliate
180 populations were starved overnight and density was determined by counting live cells. We added these
181 starved ciliates to both evolved bacterial lines. Three replicates per treatment were cultivated for 48
182 hours after which ciliate growth rate was estimated. For these final counts, we took pictures from
183 samples fixed with Lugol's solution using inverse light microscopy.

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185 **(e) Microcosm experiments with manipulated community structure**

186 Microcosms for experimental lines were set up in deep 96-well plates filled with 500 µl medium
187 containing M9 salts and 1% KB (0.2 g L⁻¹ Peptone number 3 and 0.1 ml⁻¹ glycerol). Communities

188 consisting of *P. fluorescens* and *E. coli* were assembled and either i) no ciliates (control), ii) naïve
189 ciliates or iii) coevolved ciliates were added to the microcosms. For *P. fluorescens*, we initiated
190 populations with three different levels of initial genetic diversity; i) minimal level, obtained by culturing
191 a population from a single ancestral clone, ii) full-diversity established from populations obtained from
192 the LTPE, and iii) high-diversity around ancestral trait mean. For *E. coli*, we used only one ancestral
193 clone to set up the populations. For ancestral *P. fluorescens* populations, we added one of four clones
194 isolated from the ancestor to each replicate. For the full-diversity populations, bacteria that had evolved
195 in medium either alone or together with a ciliate, both from the LTPE, were used. For the high-diversity
196 populations, we randomly combined 10 out of 20 clones representing the trait space with respect to
197 growth capacity and defence level (Fig. S2). Both species were added in even densities based on optical
198 density. The nine different treatments, consisting of three ciliates and three genetic diversity levels,
199 were replicated four times. Plates were incubated at 28 °C under constant shaking (50 r.p.m.). Every 48
200 hours, 10% of the community was transferred to fresh medium. We recorded biomass by measuring
201 optical density at 600 nm (Tecan Infinite M200 plate reader) and stored samples at –80 °C after each
202 round to archive the time series. The experimental period was 16 days, representing approximately 50
203 bacterial and ciliate generations. After reviving the archived samples from days 0, 2, 6 and 16, we
204 determined the ratio between *E. coli* and *P. fluorescens* using selective TBX and CFC agar plates,
205 respectively. With these selective media and culture conditions, we were able to clearly distinguish and
206 enumerate both bacterial species from mixed samples. From the last time point, we also isolated 10
207 individual colonies (clones) from both species that were stored at –80 °C for later analysis.

208

209 (f) Data analysis

210 All analyses were performed in R [39]. We used generalized estimating equation models (geeGLMs)
211 to compare the proportion of *P. fluorescens*, accounting for the time series structure following
212 individual microcosms. We modelled the proportion of *P. fluorescens* using `genetic diversity` and
213 `predation` both in interaction as explanatory variables together with `time` as continuous variable. To
214 account for the temporal replication, `microcosm ID` was included as a random effect. We used a model
215 of the binomial family and included an `ar1` correlation structure (continuous-time first-order

216 autoregressive correlation structure) to account for temporal correlation. We used the function *geeglm*
217 from the package *geepack* [40] with the family 'binomial' with a logit link. To model the bacterial
218 biomass data, we followed a similar approach and used estimated equation models based on the gaussian
219 family. Again, the model investigated the main effects 'genetic diversity' and 'predation' over 'time',
220 including 'ID' and an 'ar1' correlation structure. We simplified the model by dropping non-significant
221 terms. For analysis of ciliate densities, we compared only the treatments containing ciliates using a
222 model based on the gaussian family, after log10 transformation of the data. For the main effects model,
223 'genetic diversity' and 'predation' were included together with 'time', 'ID' to account for temporal
224 replication and the 'ar1' correlation structure. Again, the model was simplified to remove non-
225 significant terms. For statistical analysis, we removed day 0 measurements from all data, as these
226 represent diluted, not maximum densities of established populations represented by all the other
227 sampling points. For comparing performance of the two coevolved ciliate lines growing on the two
228 bacteria from the LTPE, we used ANOVA with ciliate ID (ciliates coevolved with *P. fluorescens* or
229 with *E. coli*) and bacterial ID (*P. fluorescens* and *E. coli*) as explanatory variables. We applied a model
230 selection process in which the interaction between both variables was dropped. To test for the effect of
231 the treatments on individual traits (defence or growth), we used generalized least squares models (gls),
232 as implemented in the *nlme* [41] package, assuming a residual variance structure dependent on the
233 experimental treatments. Multiple comparisons for gls models were performed using the package
234 *emmeans* [42], with *p*-value adjustment according to the Tukey post-hoc method and significance level
235 $\alpha = 0.05$.

236

237 3. Results

238 Our experiments suggested that the coexistence of *Escherichia coli* and *Pseudomonas fluorescens*
239 depended on both ciliate predation and genetic diversity (Fig. 1). Ciliate predation was an important
240 factor increasing *P. fluorescens* proportions in general (Table 1). Genetic diversity enhanced the
241 competitive ability of *P. fluorescens* mainly in interaction with predation. While the ancestral *P.*
242 *fluorescens* was almost completely excluded under competition, *P. fluorescens* dominated when genetic
243 diversity and predation acted together. The effect of diversity was time dependent suggesting that

244 competitive ability successively increased over the course of the experiment. The proportional changes
245 over time revealed interesting temporal dynamics congruent with the findings described above. The
246 ancestral clone was inferior to *E. coli* and experienced rapid competitive exclusion over time without
247 ciliate predation. In contrast, in the full-diversity treatment with coevolved ciliates, *P. fluorescens* was
248 the superior competitor and almost excluded *E. coli*. Adding naïve ciliates instead of coevolved ciliates
249 decreased the competitive ability of *P. fluorescens*, and both competitors seemed equal. Without
250 ciliates, *P. fluorescens* initially decreased in proportion but was able to recover when full-diversity was
251 present in the population. Interestingly, the coevolved ciliate population shifted competitive balance
252 towards *P. fluorescens* even in the absence of diversity.

253 The total bacterial biomass was independent from the diversity of *P. fluorescens* (Fig. 2), but was
254 affected by predation (Table S1). As expected, without predation, bacterial biomass was increased with
255 no difference between *P. fluorescens* diversity treatments. While there was no obvious change over
256 time under coevolved predation, bacterial biomass slightly increased over time with naïve predators. In
257 general, however, bacterial biomass seemed relatively stable over the duration of the experiment.

258 The densities of the predatory ciliates depended on total bacterial density but changed
259 independent from *P. fluorescens* diversity (Fig. 2). The ciliate densities peaked initially but decreased
260 thereafter over time resulting in low final densities (Table S2). The naïve ciliate growing on ancestral
261 *P. fluorescens* and *E. coli* had the lowest densities; however, unlike the other treatments, there was no
262 decrease over time.

263 In the bacterial clones isolated from the end-point of the experiment, there was an overall negative
264 correlation between growth and anti-predatory defence level (Pearson $r=-0.175$, $t=-3.7012$, d.f.=435,
265 $p=0.0002$) in line with a fitness trade-off between the two traits. Initial variability in the genetic diversity
266 of *P. fluorescens* and the evolutionary state of the predator together drove the *E. coli* competitor (full-
267 diversity *P. fluorescens* combined with coevolved predator), but not *P. fluorescens* itself, to diverge in
268 trait space during the experiment (Fig. 3). In the treatment with the ancestral *P. fluorescens* strain, the
269 presence of coevolved compared to naïve predators caused increased selection for anti-predatory
270 defence with a corresponding decrease in growth ability, while the opposite occurred in the full-
271 diversity treatment (for gls model results for growth and defence, and multiple contrasts, see

272 Supplementary Tables S3 and S4). This is consistent with prior coevolution in the respective predator
273 and prey populations. In turn, *E. coli* divergence in trait space was driven by increased defence and
274 decreased growth with more diverse competitors or, similar to ancestral *P. fluorescens*, coevolved
275 predators (for gls model results for growth and defence, and multiple contrasts, see Supplementary
276 Tables S5 and S6). Therefore, both predator coevolution and the genetic diversity of an otherwise
277 inferior competitor resulted in decreased resource use evolution and an increased fitness advantage of
278 anti-predatory defence.

279 As there was an effect of the different predation treatments, we also investigated whether the two initial
280 coevolved ciliate populations had different grazing capacities on the evolved bacterial lines. In all
281 conditions, we found ciliates had increased growth on *E. coli*, explaining why the inferior competitor
282 *P. fluorescens* can coexist under predation. Ciliates coevolved with *E. coli* had higher growth rates
283 compared to ciliates coevolved with *P. fluorescens* (ANOVA, $F = 7.995$, d.f.=1,9, $p = 0.0198$; Fig. S1,
284 Table S7). In turn, the ciliates were able to grow marginally better on the evolved *E. coli* bacteria
285 compared to the performance on *P. fluorescens* bacteria (ANOVA, $F = 5.772$, d.f.=1,9, $p = 0.0397$).

286

287 4. Discussion

288 Predation mediated co-existence of competitors is a widely accepted phenomenon in the field of
289 ecology. However, very little is known about how contemporary evolution and coevolution may alter
290 the operation of this mechanism. Our data provide compelling evidence for the role of genetic diversity
291 in species coexistence. While monoclonal *P. fluorescens* is rapidly outcompeted by *E. coli*, it will stably
292 coexist if the *P. fluorescens* population is genetically diverse (Fig. 1). The ensuing reduction in the
293 population size of the competitor might also alter its evolutionary dynamics, constraining resource use
294 evolution and making anti-predatory defence critical for population survival. As a result, coexistence is
295 promoted, and the genetically diverse population dominates the bacterial community. This is congruent
296 with recent theory, which predicts coexistence of diverse communities under sufficiently high trait
297 adaptation [7], and helps to explain why natural food webs contain many co-occurring species [44,45].
298 Interestingly, total biomass production seems to be independent from underlying population structure.

299 While there is obvious change in the proportions of bacterial species (Fig. 1), total bacterial and ciliate
300 production is not affected (Fig. 2). Taken together, these results indicate that the success of species in
301 communities depends on genetic variation in the traits under selection, although overall production
302 might remain unaltered which is in line with previous findings [46]. Higher biomass production would
303 be plausible, especially if both competitors only share a small resource pool. *P. fluorescens* and *E. coli*
304 should both at least slightly differ in resource use, and thus they are expected to introduce additional
305 ecological functions when both are found together [25]. However, it is possible that these functions are
306 redundant in the experimental conditions used where competitive interactions can be strong (rapid
307 outcompetition of *P. fluorescens* by *E. coli*), indicating exploitation of a similar set of resources [21,47].
308 We found predation to be highly important as an ecological force affecting coexistence of the two
309 bacteria. This result is in line with previous studies showing the effect of predation on the coexistence
310 of species [21,37,44,45,48]. A naïve predator equalizes species proportions [21], although this also
311 depends on growth-defence trade-offs [19,46]. While *E. coli* seems to grow faster in our experiment, it
312 also experiences higher loss due to predation (Fig. S1), which might explain how coexistence is
313 possible, as *P. fluorescens* seems better defended against predation loss. However, a coevolved predator
314 population which was previously exposed to different bacterial species promotes coexistence, giving
315 an advantage to *P. fluorescens*. This finding might be partly explained by the fact that *E. coli* is more
316 affected by predation and more efficient co-evolved predators might enhance this. When *P. fluorescens*
317 diversity and a coevolved predator both come together, this seems to have the strongest effect. Our
318 findings are in line with recent studies using the same experimental system, without the species
319 competition aspect, which have shown that the role of predator-prey coevolution can be an important
320 factor determining intraspecific prey diversity and eco-evolutionary feedback loops [19,43]. In addition,
321 the relative importance of ecology and evolution for co-existence has been observed to depend on the
322 community structure and the type of consumers [21].

323 In our present study, we manipulated the initial trait variation in the inferior competitor (*P.*
324 *fluorescens*). By doing this we aimed to manipulate the strength of the eco-evolutionary feedback and
325 the speed of the trait evolution by providing different amounts of initial genetic variability. We
326 hypothesised that with more within-species trait variation, it is possible that evolution is facilitated by

327 selection acting on standing genetic variation already present in the population [4,18] and population
328 trait means can quickly shift as predicted by theory [3] ultimately even changing the competitive ranking
329 between the species. Our findings support this idea, and when looking at the traits in the end of the
330 experiment, we observed that not only the co-existence of our prey species but also the trait evolution
331 of the competitor was affected by the *P. fluorescens* pre-adaptation treatment (Fig. 3d-e). Furthermore,
332 also the co-evolutionary history of the predator altered the final traits in *E. coli*, indicating that eco-
333 evolutionary mechanisms altering the co-existence of competitors constitute a process functioning in
334 different trophic levels.

335 Taken together, initial trait variability, ecological dynamics and further trait evolution are
336 interconnected processes which need to be investigated together to fully understand the role of evolution
337 in species coexistence. It is not mechanistically completely clear how traits at the end of the experiment
338 and ecological dynamic are connected in our study since links are potentially very complex. We propose
339 that understanding these feedbacks between ecological dynamics and potentially reciprocal trait
340 changes between competitors is important for our understanding of the contribution of evolution to
341 species coexistence as the focus has traditionally been mostly on ecological factors. Furthermore, we
342 need to address the role of evolution in competitors as well as the role of co-evolution in consumers.
343 When inspecting the coexistence of microbes, a recent study proposes that assembly rules in microbial
344 communities can be predicted from two- and three-way interactions for more diverse communities [24].
345 In such a model, ecological interaction is the driving force, which seems applicable as long as these
346 interactions are stable and do not change. Here we find standing genetic variation to be important at
347 both the prey and the predator level and show how it can contribute to completely shifting the
348 coexistence ratio between competitors, in turn, altering further trait change. Our study proposes that
349 evolutionary aspects cannot be neglected as they might affect interactions and therefore alter
350 coexistence. If genetic variation drives evolution, the initially estimated interaction might rapidly
351 change and depend on further interactions in a more complex community. Finally, we argue that further
352 experimental studies are needed to understand eco-evolutionary community dynamics in more species
353 rich systems such as the one we present here. Findings from these relatively simple and “unnatural”

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- 354 systems are still vital for providing mechanistic understanding on how ecological and evolutionary
355 dynamics interact in more complex natural systems.

356 **Data accessibility.** The data is available in Dryad: doi:10.5061/dryad.4b866r7

357 **Author Contributions.** T.H., L.B. and T.S. designed research; T.S. performed the experiments; J.C.
358 managed maintenance of LTPE lines; T.S., J.C. and T.H. analysed data; T.S. and T.H. wrote the first
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491 **Figure legends**

492 Figure 1. Proportion of the *P. fluorescens* population over time. The rows represent the different
493 bacterial population structures (rows in figure) for *P. fluorescens* (a–c = full-diversity; d–f = high-
494 diversity and g–i = ancestral *P. fluorescens* strain). The columns represent the different predator
495 treatments that were also applied (columns in figure, a,d,g = ciliates coevolved with bacteria, b,e,h =
496 naïve ciliate and c,f,i = no ciliates). The black line represents the proportion of *P. fluorescens* (mean \pm
497 s.e.), and the red line shows the equal proportion line as reference.

498

499 Figure 2. Total densities for bacteria and ciliates. Rows represent data from the three different
500 population structures (a–c = full-diversity; d–f = high-diversity and g–i = ancestral *P. fluorescens* strain
501 without diversity) and columns represent the three different predation treatments (columns in figure,
502 a,d,g = ciliates coevolved with bacteria, b,e,h = naïve ciliate and c,f,i = no ciliates). Orange lines and
503 points (mean \pm s.e.) show total bacterial density measured by absorbance, and blue lines and squares
504 represent ciliate densities (mean \pm s.e.). Bacterial density is shown as optical density at 600 nm, and
505 ciliate density as cells ml⁻¹, both normalized to 0–1 range.

506

507 Figure 3. Divergence in trait space caused by the genetic diversity of *P. fluorescens* or predator
508 evolutionary history (ellipses depict 95 % confidence levels). Growth is the biomass yield in the absence
509 of predation and defence is the effect of predation (0 = predation has no effect) on the biomass yield
510 (both in optical density area units). Panels a-c, *P. fluorescens*, and panels d-f, *E. coli* clones isolated
511 from the end point of a microcosm experiment.

512

513

514 **Table 1.** A Generalized Estimated Equation model showed that the main effects `predation` and
 515 `diversity` had both significant effects on *Pseudomonas fluorescens* proportions
 516

	Df	X2	P(> Chi)	
Time	1	5	0.0259	*
Diversity	2	5.4	0.0689	.
Predation	2	142	0.0000	***
Time x Diversity	2	6.1	0.0481	*
Time x Predation	2	13	0.0015	**
Diversity x Predation	4	25.4	0.0000	***
Time x Diversity x Predation	4	22.7	0.0001	***

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

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