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AsChip: A High-Throughput qPCR Chip for Comprehensive Profiling of Genes Linked to Microbial Cycling of Arsenic

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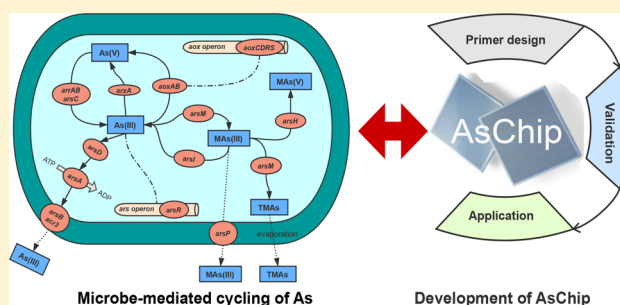
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Supporting Information

ABSTRACT: Arsenic (As) is a ubiquitous toxic element adversely affecting human health. Microbe-mediated cycling of As is largely mediated by detoxification and energy metabolism in microorganisms. We here report the development of a novel high-throughput qPCR (HT-qPCR) chip (AsChip) for comprehensive profiling of genes involved in microbial As cycling (here collectively termed “As genes”). AsChip contained 81 primer sets targeting 19 As genes and the 16S rRNA gene as a reference gene. Gene amplicon sequencing showed high identity (>96%) of newly designed primers corresponding to their targets. AsChip displayed high sensitivity (plasmid template serial dilution test; $r = -0.99$), with more than 96% of all PCR assays yielding true positive signals. R^2 coefficients for standard curves and PCR amplification efficiencies averaged 0.98 and 0.99, respectively. A high correlation between C_T values obtained by AsChip and conventional qPCR was obtained ($r = 0.962$, $P < 0.001$). Finally, we successfully applied AsChip on soil samples from a chromium–copper–arsenic-contaminated field site and identified diverse As genes with total abundance average of 0.4 As gene copies per 16S rRNA. Our results indicate that AsChip constitutes a robust tool for comprehensive quantitative profiling of As genes in environmental samples.



INTRODUCTION

Arsenic (As) is a ubiquitous toxic element impacting human health.¹ It is present in the atmosphere, hydrosphere, pedosphere, and biosphere with both natural and anthropogenic origins.^{2–5} Arsenic in the environment primarily poses health risks to humans via water⁶ or food^{7–9} and has been linked to various cancers,¹⁰ cardiovascular disease,¹¹ diabetes,¹² and skin lesions.¹³ The toxicity, mobility, and environmental fate of As are to a large extent determined by its chemical speciation.¹⁴ For example, inorganic As species prevalent in drinking water are much more toxic than arsenobetaine, which is the primary source of As in fish meat.¹⁵ Hence, the following toxicity order of arsenicals have been reported for humans: inorganic As(III) > dimethylarsenite [DMAs(III)] > dimethylarsenate [DMAs(V)] > monomethylarsenate [MMAs(V)] > inorganic As(V).¹⁶

Environmental As speciation is profoundly impacted by microbial activity. Dissimilatory As oxyanion oxidation and reduction constitute ancient types of bacterial energy

metabolism,¹⁷ whereas other As biotransformation processes can be linked to As detoxification processes conferring bacterial resistance to As.^{18,19} Collectively, microorganisms (mainly prokaryotes) transforming As can be divided into several more or less distinct functional groups,^{2,3,20} such as chemoautotrophic arsenite-oxidizing bacteria (CAO) generating energy from the dissimilatory oxidation of arsenite with O₂ or nitrate as terminal electron acceptor, dissimilatory arsenate-reducing anaerobic prokaryotes (DARP) generating energy from the oxidation of various organic or inorganic electron donors with arsenate as the terminal electron acceptor, and prokaryotes transforming As compounds for other physiological purposes (e.g., As resistance/detoxification). Where the two first mentioned groups consist of specialized prokaryotes

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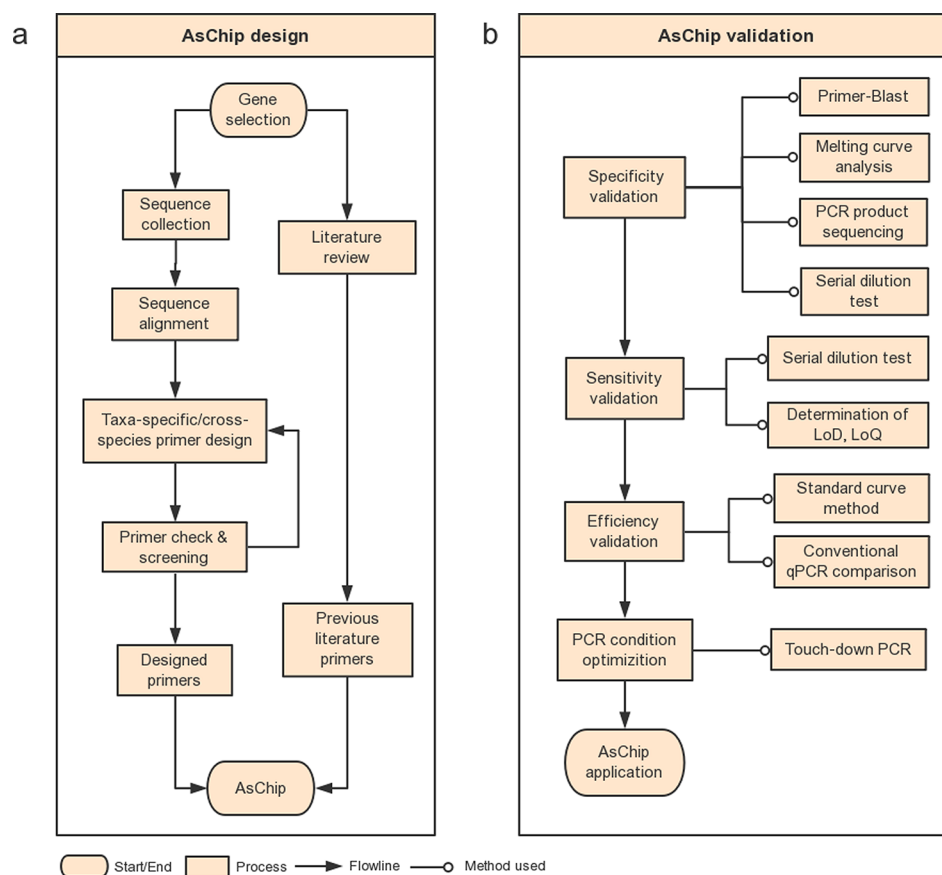


Figure 1. Flowchart of major steps for development of AsChip, including (a) design and (b) validation.

generating energy from inorganic As biotransformations, the latter group includes prokaryotes carrying out a wide range of contrasting As biotransformation processes. A major subgroup includes heterotrophic arsenite oxidizers (HAO) transforming toxic arsenite into the less toxic arsenate. Conversely, other bacteria reduce arsenate into arsenite and couple it to arsenite efflux to confer As resistance. Finally, a range of microorganisms transform organoarsenicals, of which many are volatile and highly mobile. Despite their immense biogeochemical significance, only little is known about the physiological purposes of these As biotransformations.²

A large number of genes (here collectively conferred to as “As genes”) are involved in microbe-mediated cycling of As and constitute promising functional gene markers for studies of As biotransformation processes and the various functional groups of prokaryotes that carry out these processes.¹⁸ For example, inorganic arsenate can be reduced to arsenite by arsenate reductase encoded by *arsC* (resistance) or *arrAB* (anaerobic As respiration), whereas the toxic As(III)/MAs(III) can be pumped out of bacterial cells by arsenic transporters encoded by *arsB*, *arsP*, or *acr3* (resistance). Arsenite or organoarsenical oxidases encoded by *arxA* (dissimilatory arsenite oxidation), *aioAB* (resistance), or *arsH* (resistance) may further transform As(III)/MAs(III) to the less toxic As(V)/MAs(V). As methylation is encoded by *arsM* (resistance) and results in the loss of As by volatilization or products such as mono-, di- or trimethylarsines,²¹ demethylation of trivalent organoarsenicals (MAs[III] and roxarsone) is catalyzed by C-As lyase encoded by *arsI* as part of a resistance pathway.^{22–24}

Our knowledge about As biotransformation processes and the underlying microbial ecology is incomplete in part due to the lack of comprehensive, targeted methods to comprehensively study As-transforming microorganisms. With the development of next-generation sequencing (NGS) technologies and their decreasing costs, metagenomic and metatranscriptomic studies of As biotransformation are beginning to emerge.^{25–27} However, these approaches only provide relative abundances of targeted genes and their transcripts and should therefore ideally be complemented by their high throughput quantification. Quantitative polymerase chain reaction (qPCR) represents one of the most widely applied techniques for quantifying functional genes and has already been used in several environmental studies of As biogeochemistry.^{21,28–37} This technique generally provides more sensitive detection of genes in environmental samples as compared to hybridization microarray technology (e.g., the GeoChip).^{38,39} However, conventional single-gene qPCR gets laborious and impractical when many genes have to be quantified at the same time as would be necessary for comprehensive monitoring of all As genes in environmental samples.^{35,37} Hence, we aimed to develop and validate a novel high-throughput qPCR (HT-qPCR)-based chip for sensitive, rapid, and cost-efficient quantification of known prokaryotic As genes. The developed HT-qPCR chip (AsChip) targets 19 microbe-mediated As cycling genes and the bacterial 16S rRNA gene as a reference gene. AsChip employs a total of 81 cross-species primer sets covering hundreds of species with 70 primer sets being newly developed and validated as part of the present study.

Table 1. Summary of Genes Targeted by AsChip, Microorganisms Hosting Genes, and Numbers of AsChip qPCR Primers Used To Detect Genes^a

Gene type	Functional process	Encoded proteins	Functionality	Microbes	No. of primer pairs		ref
					Designed	Published	
<i>aoxA/aioB</i>	As(III) oxidation	As(III) oxidase small subunit	As(III) oxidation	HAO	3	1	73
<i>aoxB/aioA</i>		As(III) oxidase large subunit	As(III) oxidation	HAO	3	1	73
<i>aoxR/aioR</i>		response regulator	regulation of <i>aox/aio</i> operon expression	HAO	3	1	73
<i>aoxS/aioS</i>		sensor histidine kinase	part of two-component signal transduction system	HAO	3	1	73
<i>arsH</i>		organoarsenical oxidase	trivalent organoarsenicals oxidation	ARM	3	0	23
<i>aoxC/aioC</i>		cytochrome c	encoding c-type cytochrome in <i>aox/aio</i> operon	HAO	1	1	74
<i>aoxD/aioD</i>		chIE - molybdoptein biosynthesis	encoding an enzyme involved in molybdoptein biosynthesis in <i>aox/aio</i> operon	HAO	1	1	74
<i>arxA</i>		As(III) oxidase	As(III) oxidation	CAO	1	0	75
<i>arrA</i>	As(V) reduction	As(V) respiratory reductase large subunit	As(V) reduction	DARP	1	1	36, 76
<i>arrB</i>		As(V) respiratory reductase small subunit	As(V) reduction	DARP	1	0	36, 77
<i>arsC</i>		As(V) reductase	As(V) reduction	ARM	13	2	77
<i>arsR</i>		transcriptional repressor	regulation of <i>ars</i> operon expression	ARM	8	0	77
<i>arsI</i>	Arsenic methylation and demethylation	C-As bond lyase	demethylation of trivalent organoarsenicals to less toxic As(III)	AMM	4	0	23
<i>arsM</i>		As(III) methyltransferases	arsenic methylation	AMM	5	0	23
<i>arsA</i>	Arsenic transport	As(III)-activated ATPase	catalysis of oxyanion-translocating ATPase	ARM	7	0	78
<i>arsB</i>		As(III) efflux pump protein	extrusion of As(III) from the cell	ARM	6	0	79
<i>arsD</i>		arsenical metallochaperone	transfer of trivalent metalloids to <i>arsA</i>	ARM	3	0	78
<i>arsP</i>		trivalent organoarsenicals permease	extrusion of trivalent organoarsenicals from the cell	ARM	2	0	80
<i>acr3</i>		As(III) efflux pump protein	extrusion of As(III) from the cell	ARM	2	1	81, 82
Total 19 genes					70	10	

^aDetailed information on all primer sets is shown in Table S1. Abbreviations used: Heterotrophic arsenite oxidizers, HAO; Arsenic resistant microbes, ARM; Chemoautotrophic arsenite oxidizers, CAO; Dissimilatory arsenate-reducing prokaryotes, DARP; Arsenic methylating microbes, AMM.

MATERIALS AND METHODS

Design of AsChip. Major steps for AsChip design are shown in Figure 1a. A total of 19 targeted genes of relevance for microbial cycling of As were selected (Table 1). Collectively, these targeted genes encode As(III) oxidation, As(V) reduction, As methylation, As demethylation, and As transport (efflux). Ten validated primer pairs were selected from previously published studies, whereas a total of 70 novel primer pairs were designed as part of this study (Table 1; Table S1). Primer design was performed based on target gene sequences retrieved from the National Center for Biotechnology Information (NCBI) GenBank database. Gene sequences were obtained from both bacteria and archaea and were further collected and aligned using MEGA 7.0 with ClustalW. Sequences were clustered into different groups using a neighbor-joining phylogeny tree based on alignment results. SYBR Green Design of AlleleID 6.0 was used to design tax-specific/cross-species primers for top clustered groups of each gene. The cross-species primers were designed based on conserved regions found in clustered groups. The quality of designed primers and expected priming efficiency were further

checked by AlleleID 6.0 with the following criteria: primer length of 18–24 bases, amplicon length of 50–200 bases, GC content of 40%–60%, expected melting temperature (T_m) of product at 48–62 °C, hairpin $\Delta G \geq -3$ kcal/mol, self-dimer $\Delta G \geq -6$ kcal/mol, cross-dimer $\Delta G \geq -6$ kcal/mol, and GC clamp ≤ 3 .^{40,41} Only primers meeting the above quality criteria were used for further validation and application. The bacterial 16S rRNA gene was included in the construction of AsChip as a reference gene using the universal primers 1108F/1132R.⁴² Detailed information on the 81 primer sets is shown in Table S1.

Validation of AsChip. Major steps of AsChip validation are shown in Figure 1b. The specificity of AsChip was validated computationally and experimentally. Computationally, Primer-BLAST⁴³ was used to check newly designed primers for specificity against the NCBI nucleotide collection (nr) database with bacteria (taxid:2) and archaea (taxid:2157) as target organisms with a max target size of 1000 bases. All newly designed primers were also tested experimentally by PCR using DNA extracted from contrasting environmental samples as templates. DNA were extracted from samples

collected from a chromium–copper–arsenic contaminated soil,⁴⁴ estuaries (sediment),⁴² and urban sewage (water).⁴⁵ Water samples were filtered through a 0.22 μm cellulose nitrate membrane filter to collect bacterial pellets prior to DNA extraction. Detailed protocols for DNA extraction of sediment and water samples were described previously.^{42,45} The detailed protocol for DNA extraction from soil samples is described in the section entitled [Application of AsChip on Environmental Samples](#). The purity of DNA was checked (ND-1000, Nanodrop, USA), and DNA was diluted to the same concentration with sterile water for later use. A total of 53 different PCR amplicons were detected and checked via agarose gel electrophoresis and Sanger sequencing⁴⁶ (company Bioray, Xiamen, China). Amplicon sequences were assembled and aligned with target sequences using the software DNAMAN v9.0. Furthermore, AsChip primer specificities were routinely monitored by analysis of melting curves in all downstream AsChip HT-qPCR experiments using the Wafergen SmartChip Real-time PCR system. For regular PCR, each 50 μL PCR reaction contained 25 μL TaKaRa PremixTaq (ExTaq), 1 μL of each primer, 22 μL PCR grade water, and 1 μL DNA template. The thermal cycle consisted of an initial polymerase activation at 95 $^{\circ}\text{C}$ for 10 min and 40 cycles of denaturation (95 $^{\circ}\text{C}$, 30s), annealing (58 $^{\circ}\text{C}$, 30s), and extension (72 $^{\circ}\text{C}$, 120s).

The serial dilution test and standard curve analysis were conducted to validate specificity, sensitivity, and efficiency using plasmids as positive controls. Gene fragments targeted by all the 70 newly designed primers were synthesized and cloned into plasmids cloning vector pUC57-Ampicillin (GENEWIZ, Inc.). The size and sequence of all the artificial target fragments (5' \rightarrow 3') are listed in [Table S1](#). To develop standard curves for further evaluation, these plasmid DNA containing target gene fragments were mixed (40 ng each) into one mixed positive control. Ten-fold dilution series covering from 1 to 10^{11} plasmid copies per reaction were prepared to generate standard curves for AsChip HT-qPCR analyses. Standard curve fitting was performed with linear regression analysis using OriginPro 9.1. The optimal cutoff point was determined as the maximum qPCR cycle of quantitative detection in serial dilution with true-positive and true-negative signals. The limit of detection (LoD) was determined as the lowest amount of target that could be detected. The limit of quantification was determined as the lowest amount of target that could be quantified with true-positive/negative signal, corresponding to the optimal cutoff point in the standard curve. The qPCR primer amplification efficiency was calculated using standard curve with the equation: $\text{Eff} = 10^{(-1/\text{slope})} - 1$, where slope corresponds to the slope of the standard curve. AsChip HT-qPCR data generated for all 70 novel primer pairs was further validated by conventional qPCR conducted on the Light-Cycler480 Real-Time PCR System using the same PCR conditions and the same environmental samples.

The qPCR conditions for the AsChip HT-qPCR system were optimized using touch-down PCR (TD-PCR) with five randomly selected designed primer pairs.⁴⁷ TD-PCR products were checked by agarose gel electrophoresis. An optimal annealing temperature of 58 $^{\circ}\text{C}$ was determined by TD-PCR with the expected optimum annealing temperature (T_a Opt) and melting temperature (T_m) of the product generated by AlleleID also considered. The thermal cycle consisted of initial enzyme activation 10 min at 95 $^{\circ}\text{C}$, followed by 40 cycles of denaturation at 95 $^{\circ}\text{C}$ for 30s, annealing at 48 to 60 $^{\circ}\text{C}$ with

temperature increments of 0.3 $^{\circ}\text{C}$ per cycle for 30s, and extension at 72 $^{\circ}\text{C}$ for 120s. Apart from this, TD-PCR reaction conditions were the same as described previously for regular PCR. Template DNA was extracted from the CCA contaminated soil site in Fredensborg as described below. All regular PCR/TD-PCR were performed on Bioer LifePro Thermal Cycler (Bioer Technology Co., Ltd.) with technical triplicates and included negative controls.

Application of AsChip on Environmental Samples.

AsChip was tested with DNA extracted from three topsoil samples from an abandoned Danish wood impregnation site contaminated with different levels of chromium, copper, and arsenic.^{44,48} Approximately 500 g of three topsoil samples were collected in sterile plastic bags. After homogenization, approximately 5 g of soil was transferred to sterile 15 mL tubes, transported to the laboratory (<1 h of transport) in an ice-cooled box, and subsequently stored at -20°C until DNA extraction. A MoBio PowerSoil DNA Isolation Kit was used for DNA extraction. DNA quality was checked via ultraviolet absorbance (ND-1000 Spectrophotometer, NanoDrop, Thermo Fisher Scientific). DNA was diluted to 30 ng μL^{-1} with sterile PCR grade water and stored at -20°C . Soil DNA was subsequently analyzed by HT-qPCR AsChip on Wafergen SmartChip Real-time PCR system with the format of 81 assay X 54 sample in 50 nL reaction system. The thermal cycle consisted of initial enzyme activation for 10 min at 95 $^{\circ}\text{C}$, followed by 40 cycles of denaturation at 95 $^{\circ}\text{C}$ for 30s, and annealing/extension at an optimized temperature of 58 $^{\circ}\text{C}$ for 30s. Only genes with positive amplification of all technical triplicates were considered positive for gene detection and subsequently used for data analysis. All HT-qPCRs performed in this study was performed in technical triplicates and included negative and positive controls.

Statistical Analysis. AsChip data were analyzed to quantify genes involved in microbial As cycling. Absolute abundance (copy numbers per gram of soil) of 16S rRNA genes was calculated with the standard curve method of quantification with a C_T value of 31 as the detection limit where C_T equals the threshold cycle.⁴⁹ Absolute As gene copy numbers were calculated using the following formula: absolute As gene copy number = (relative As gene copy number/relative 16S rRNA gene copy number) \times absolute 16S rRNA gene copy number, where relative gene copy number = $10^{(31-C_T)/(10/3)}$.⁵⁰ Normalized As gene abundances were calculated as As gene copy numbers per 16S rRNA gene. The flowchart was generated using ProcessOn (Damai Co. Ltd., Beijing, China). The bar chart, box chart, and scatter diagram were plotted by OriginPro 9.1. Linear regression analyses were performed on OriginPro 9.1, and the liner fit, confidence, and prediction bands were generated automatically. Heatmap analysis and cluster analysis were performed using HemI.⁵¹

RESULTS AND DISCUSSION

Coverage and Specificity of AsChip. The developed AsChip targets almost all known prokaryotic As genes^{2,18} ([Table 1](#) and [Table S1](#)). Validations of AsChip primer specificities were performed both *in silico* using Primer-BLAST and experimentally. Primer-BLAST identified a total of approximately 6500 target sequences with the 70 newly designed primers. Almost all gene targets (98.9%) were derived from diverse bacterial phyla, whereas the remaining fraction originated from archaea (Euryarchaeota). Bacterial phyla

covered by AsChip primers included Proteobacteria, Firmicutes, Actinobacteria, Bacteroidetes, Cyanobacteria, Verrucomicrobia, and Fusobacteria. We randomly selected five As genes in order to calculate the fraction of known annotated As gene sequences (Genbank) that could be amplified by the designed AsChip primers. The fraction of sequences targeted by the selected As genes were as follows: *aoxA* (67%; 4 out of 6), *arsC* (80%; 228 out of 286), *arsH* (60%; 116 out of 192), *arsM* (62%; 46 out of 74), and *arsR* (64%; 492 out of 772), demonstrating adequate coverage of AsChip. Primer-BLAST matches of 100% between primers and their target sequences were obtained for all designed primers using the NCBI nr database (organisms limited to Bacteria and Archaea). Primer specificities were further tested experimentally by the melting curve analysis on HT-qPCR reactions with all 70 newly designed primers targeting As genes. All melting curves of qPCR generated with AsChip showed the presence of a single peak, suggesting the success of designed primers for their specificity. The specificity was double checked and further confirmed by Sanger sequencing of all amplified PCR products ($n = 53$) obtained with newly designed primer pairs. Agarose gel electrophoresis of amplified PCR products in all cases showed a single band of the expected size, and DNA sequencing further revealed that the obtained amplicons corresponded well to their respective targets (Figure 2).

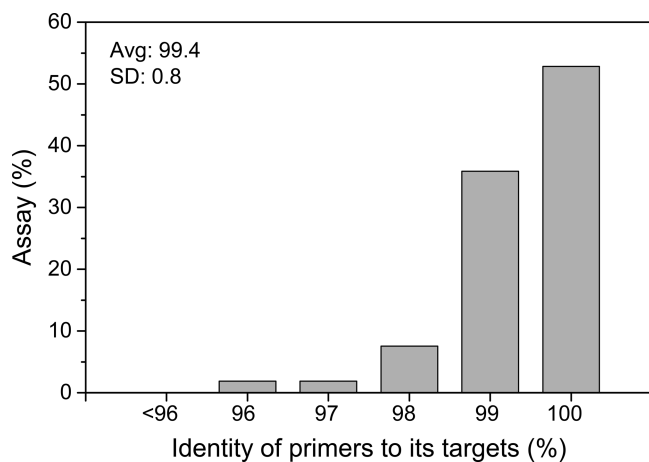


Figure 2. Specificity of newly designed AsChip primers was evaluated by sequencing of their corresponding PCR amplicons ($n = 53$). The figure depicts the distribution of PCR assays (%) against the amplicon sequence similarity (%) to their respective targets used for primer design.

Hence, amplicon sequences in all cases showed high similarity to their targets (>96%) with 89% of the primers resulting in amplicon sequences with $\geq 99\%$ similarity to their respective targets. The serial dilution test displayed true-positive amplifications of only targeted sequences, whereas no amplicons could be detected for negative controls (i.e., no template) with the absence of any primer secondary structures. Collectively, our results thus demonstrated a high and satisfactory specificity of AsChip suggesting a low risk for false positive detection of As genes.

Sensitivity and Efficiency of HT-qPCR AsChip and Comparison with Conventional qPCR. The performance of AsChip was assessed by estimating its sensitivity using standard curves generated by serial dilution test with mixed plasmids serving as positive controls (Figure 3; Figure S1).

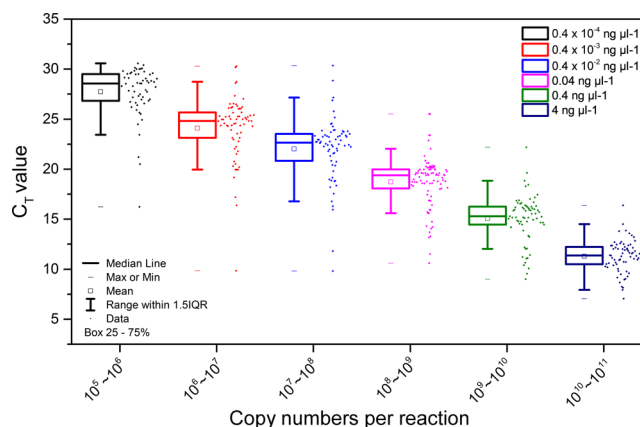


Figure 3. AsChip sensitivity (serial dilution test) with 70 newly designed primers as evaluated by 10-fold serial dilution of mixed positive plasmid controls. Dots on the right side of boxes represent data obtained for individual primer pairs. Plasmid DNA concentrations are indicated in the figure (upper right). IQR, interquartile range.

The results demonstrated a high sensitivity of AsChip. C_T values corresponding to positive plasmid template concentrations of 4×10^{-5} to $4 \text{ ng } \mu\text{L}^{-1}$ ranged from 7 to 31. Within this template concentration range, more than 96% of all PCR assays yielded true positive signals, whereas standard curve Pearson correlation coefficients and R^2 (4×10^{-5} to $4 \text{ ng } \mu\text{L}^{-1}$ range) averaged -0.99 and 0.98 , respectively (Table S2). For the more efficient application of AsChip, the optimal cutoff point, the LoD, and the limit of quantification (LoQ) were determined by standard curve analysis of serial dilution (Table S3). The optimal cutoff points for C_T values of the 70 newly designed primers ranged from 26 to 31 with an average of 30. The LoD varied from 7 to 262 target copies with 61 as the average, while LoQ varied from 11 to 132,667 target copies with an average of 10,319.

HT-qPCR AsChip performance for all 70 newly designed primers was further validated by calculating qPCR efficiency indices and R^2 based on standard curves (Figure 4; Table S2). PCR efficiency indices averaged 0.99 with a standard deviation

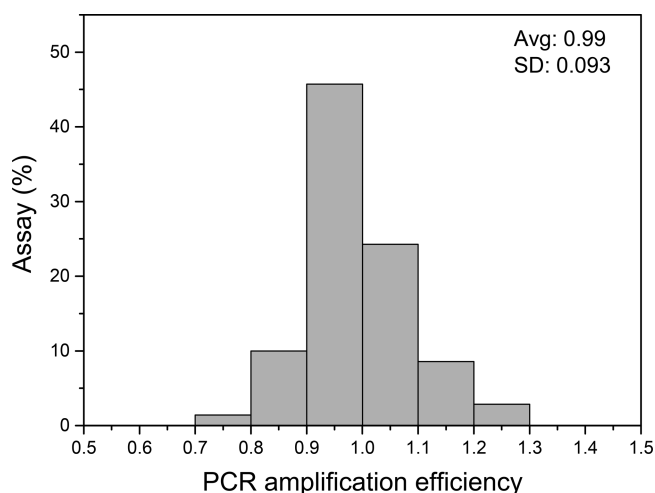


Figure 4. Distribution of AsChip HT-qPCR amplification efficiency with 70 newly designed primers. Standard curves and their parameters are shown in Figure S1 and Table S2.

of 0.09 and ranged between 0.77 and 1.26 with the vast majority in the range of 0.8–1.2 (Figure 4). R^2 coefficients for standard curves averaged 0.98 ± 0.01 (\pm standard deviation). A potential deviation on normalized abundances (i.e., As gene copy number per 16S rRNA gene) might occur when applying comparative C_T since it was based on the assumption of the same amplification efficiency of all amplicons. However, for future AsChip application, this can be overcome with absolute quantification of each target with the standard curve method, by increasing primer concentration, or by optimizing the primer to specific T_m .

The high sensitivity and robust performance of AsChip were further confirmed by comparing C_T values obtained by nanoliter-scale AsChip analysis with traditional microliter-scale qPCR (Figure 5). C_T values obtained by the two qPCR

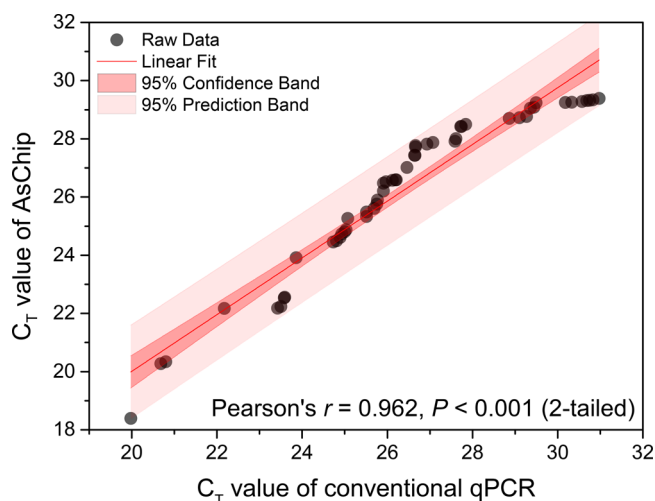


Figure 5. AsChip performance as evaluated against conventional single-gene qPCR by comparison of C_T (cycle threshold) values ($n = 52$). Linear regression analysis revealed a linear correlation (Pearson's $r = 0.962$, $P < 0.001$) between the C_T values obtained from the two qPCR platforms (conventional single-gene qPCR and AsChip).

platforms showed a significant positive correlation (Pearson's $r = 0.962$, $P < 0.01$), indicating a high degree of reliability and reproducibility of AsChip (Figure 5). Hence, similar PCR efficiencies and detection limits were obtained by the two qPCR platforms in accordance with previous reports from the development of other HT-qPCR chips.^{52,53} A total of 52 As gene targets could be detected by both conventional qPCR and AsChip, whereas an additional seven As genes could only be detected by AsChip suggesting a higher sensitivity of AsChip. Other possible explanations for the observed slight differences between results obtained with the two qPCR platforms include automated versus manual qPCR reagent handling, laboratory contaminations, and inaccuracies of DNA quality quantified by UV absorbance.

Application of AsChip. AsChip analysis was carried out on DNA extracted from three Danish top soils collected from a site contaminated with different levels of chromium, copper, and arsenic. The three soils were labeled “low”, “medium”, or “high”, depending on the arsenic contamination level (33, 203, or 2839 $\mu\text{g As g}^{-1}$, respectively). The absolute abundances of As genes (copies per gram of soil) ranged from 8.28×10^9 to 1.34×10^{11} (average of 7.30×10^{10} ; Table S4), but total normalized abundances varied much less and were 0.43, 0.51, and 0.32 in low, medium, and high contaminated soils. Hence, our data indicate that the overall As gene abundance did not vary much with soil As contamination level.

The normalized abundance of individual As genes are shown in Figure 6. Out of 19 different As genes, 17 could be detected in the three soils demonstrating high potential for microbial As transformations. Remarkably, the *arsX* gene encoding the arsenite oxidase in chemoautotrophic arsenite-oxidizing bacteria (CAO) was not detected in any of the three contaminated soils suggesting that these specialized bacteria generating energy from the oxidation of arsenite with molecular oxygen or nitrate as terminal electron acceptor were not active in the studied soils. This finding is consistent with arsenate being the dominant inorganic As species and the

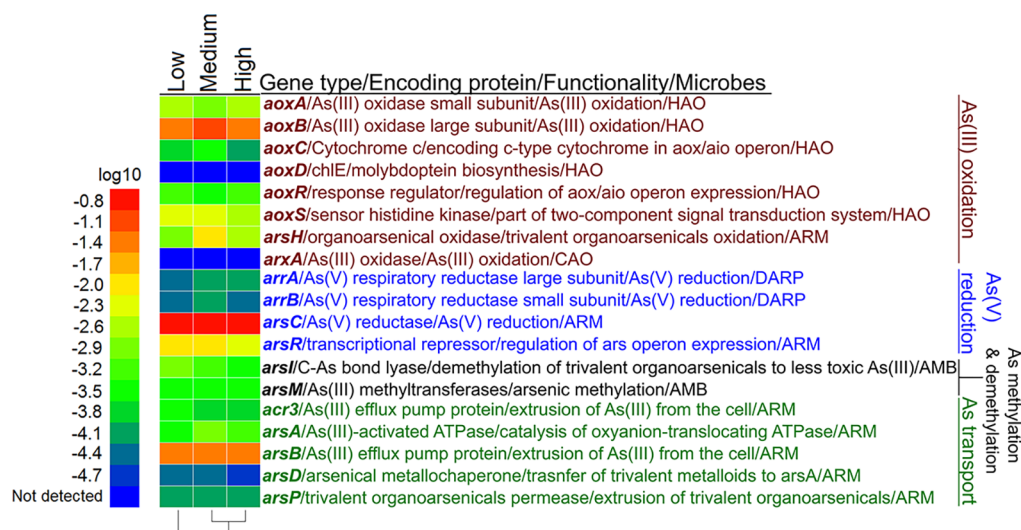


Figure 6. \log_{10} transformed normalized As gene abundance (copy number per 16S rRNA) of 19 genes of relevance for microbe-mediated biogeochemical cycling of As in three top soils. Soils were collected from a chromium–copper–arsenic-contaminated site and named “Low” (33 $\mu\text{g As g}^{-1}$), “Medium” (203 $\mu\text{g As g}^{-1}$), or “High” (2839 $\mu\text{g As g}^{-1}$). Samples were clustered based on Pearson distance using hierarchical algorithms (average linkage cluster). Abbreviations used: HAO, heterotrophic arsenite oxidizers; ARM, arsenic resistant microorganisms; CAO, chemoautotrophic arsenite oxidizers; DARP, dissimilatory arsenate-reducing prokaryotes.

anticipated absence of an As redox stratification in the studied upland soils.^{54–56}

Two of the most abundant As genes were *arsB* (0.049–0.069 copies per 16S rRNA gene) and *arsC* (0.18–0.30 copies per 16S rRNA gene). Collectively, these genes encode arsenate resistance by the arsenate reduction-arsenite efflux pathway.¹⁸ Hence, our AsChip data indicates that the *arsBC* encoded resistance mechanism is very common in bacteria irrespective of the As contamination level of their natural habitats. This is consistent with the frequent occurrence of *arsBC* in cultured bacteria (e.g., *Escherichia coli*) isolated from noncontaminated habitats.³⁰ Interestingly, AsChip also consistently detected a series of As genes (*arsH*, *arsI*, and *arsM*) involved in recently discovered parallel pathways for organoarsenical biotransformation and detoxification.²⁵ The *arsM* gene encodes a As(III) S-adenosylmethionine methyltransferase that transforms inorganic arsenite into the even more toxic organoarsenical methyl arsenite (a human carcinogen) which can be rapidly oxidized to its less toxic pentavalent form, whereas *arsI* and *arsH* encode different methyl arsenite detoxification mechanisms.^{23,57} Other arsenic-resistant genes (*arsADRP*, *acr3*) were also detected in our samples demonstrating the existence of several parallel As resistance pathways in the studied microbial communities.

*AoxABC*RS genes (also called *aioABC*RS)¹⁸ are involved in heterotrophic arsenite oxidation and were detected in all samples with the *aoxB* gene being the most abundant of the *aox* genes and the second most abundant As gene overall (0.04, 0.11, and 0.07 copies per 16S rRNA gene in low, medium, and high contaminated soil, respectively). The *aoxB* gene (large Mo-protein subunit of As(III) oxidase at 7.1×10^9 , 2.8×10^{10} , and 1.7×10^9 copies g⁻¹, correspondingly) was 10, 45 and 26 folds more abundant than the *aoxA* gene (small Fe-S Rieske subunit of As(III) oxidase, 5.7×10^8 , 6.3×10^8 , and 6.7×10^7 copies g⁻¹, correspondingly) in the soils named low, medium, and high, respectively (Table S4).⁵⁸ This was not due to differences in PCR amplification efficiencies and could thus most likely be attributed to lower coverage of the primers used for *aoxA* as compared to *aoxB*. This interpretation is consistent with *aoxB* (also named *aioA*) being more frequently studied as compared to *aoxA* (also named *aioB*), and more target sequences were thus available in databases for *aoxB* primer design. Nevertheless, differences in the normalized abundance for the three studied soils followed the same order for both *aoxA* and *aoxB* suggesting that the current first-generation AsChip can still be used to monitor relative differences between samples. In contrast to the *aoxAB* gene pair, the genes *arrA* and *arrB* encoding the large and the small subunit of As(V) reductase in dissimilatory arsenate-reducing prokaryotes (DARPs) were found in near equal abundances for each soil, suggesting that AsChip primers used for detection of *arrA* and *arrB* displayed similar bacterial community coverage. Hence, *arrA* gene copy numbers were 6.2×10^6 , 2.4×10^7 , and 2.1×10^6 in soil named low, medium, and high, whereas corresponding numbers for *arrB* were 10.0×10^6 , 2.9×10^7 , and 1.2×10^6 copies g⁻¹, respectively (Table S4).

To our knowledge, AsChip represents the first HT-qPCR-based chip for As genes, and our results indicate that it can be used as a robust and reliable tool for comprehensive quantitative profiling of As genes in environmental samples. Hence, AsChip provides specific, sensitive, and targeted quantification of As genes and may be used in various environments for better understanding of As microbe-

mediated biotransformations and its biogeochemical cycling. The high sensitivity and PCR reaction miniaturization allow AsChip analysis to be performed with only a small amount of template DNA. Furthermore, HT-qPCR-based chips have already been successfully developed and applied for profiling other genes in various environments.^{42,59–63} Consequently, HT-qPCR data analysis is already well developed and allows researchers to perform absolute quantification,⁴² relative quantification,^{50,64} or comparative quantification^{61,65,66} of the targeted genes. Moreover, results obtained from AsChip may potentially be used to link detected As genes to their putative microbial hosts via the Primer-BLAST tool.⁴³ The microarray-based GeoChip³⁹ also includes probes targeting As genes and has been used in a previous study of As biotransformation processes.⁶⁷ However, the GeoChip was designed for a wide range of functional genes and not specifically for As genes, and only a limited diversity of As genes were targeted (i.e., *arsABC*M and *aoxB*). Compared with GeoChip, AsChip presented in this study thus displays better As gene coverage and, most likely, higher sensitivity for As gene quantification.

Although the current first-generation AsChip has great potential as a comprehensive molecular tool for studies of microbial As cycling, novel functional groups of As-biotransforming prokaryotes and As genes not matching the current AsChip primers are likely to be discovered and characterized in the future. This implies that AsChip will have to be further refined and improved. The literature already contains reports of As genes that were not included in AsChip design. For instance, two novel arsenic resistance genes *arsO* (for a putative flavin-binding monooxygenase) and *arsT* (for a putative thioredoxin reductase) were reported in a previous study.⁶⁸ Likewise, the As resistance genes *arsN* (encoding acetyltransferase),⁶⁹ *arsJ* (encoding organoarsenical efflux permease),⁷⁰ and *gstB* (encoding glutathione S-transferase B)⁷¹ were reported previously, but only limited coding sequence information is currently available for primer design. Despite these limitations, AsChip offers unprecedented potential for targeted, yet comprehensive, quantification of As genes in environmental samples. We propose that AsChip may be applied together with other “omics” methods and (geo)chemical analysis in future studies of As resistance, biotransformation, and biogeochemistry. Such polyphasic approaches may reconcile geochemical and genomics data for the development of systems models to predict the fate and impacts of As and the movement of microorganisms in the environment.^{18,72}

■ ASSOCIATED CONTENT

📄 Supporting Information

The Supporting Information is available free of charge on the ACS Publications website at DOI: 10.1021/acs.est.8b03798.

Standard curves obtained with all 70 newly designed AsChip primers (Figure S1); parameters of standard curves (Table S2); optimal cutoff point, LoD, and LoQ (Table S3); absolute abundances of the 16S rRNA gene and As genes involving different functional processes in three soils (Table S4). (PDF)

Detailed information on 81 AsChip primer sets and their targets (Table S1). (XLSX)

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Notes

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REFERENCES

- (1) Abernathy, C. O.; Thomas, D. J.; Calderon, R. L. Health effects and risk assessment of arsenic. *J. Nutr.* **2003**, *133* (5), 1536S–1538S.
- (2) Zhu, Y. G.; Yoshinaga, M.; Zhao, F. J.; Rosen, B. P. Earth abides arsenic biotransformations. *Annu. Rev. Earth Planet. Sci.* **2014**, *42*, 443–467.
- (3) Oremland, R. S.; Stolz, J. F. The ecology of arsenic. *Science* **2003**, *300* (5621), 939–944.
- (4) Chen, W. Q.; Shi, Y. L.; Wu, S. L.; Zhu, Y. G. Anthropogenic arsenic cycles: a research framework and features. *J. Cleaner Prod.* **2016**, *139*, 328–336.
- (5) Ferguson, J. F.; Gavis, J. A review of the arsenic cycle in natural waters. *Water Res.* **1972**, *6* (11), 1259–1274.
- (6) Smith, A. H.; Lingas, E. O.; Rahman, M. Contamination of drinking-water by arsenic in Bangladesh: a public health emergency. *Bull. World Health Organ.* **2000**, *78* (9), 1093–1103.
- (7) Brammer, H.; Ravenscroft, P. Arsenic in groundwater: a threat to sustainable agriculture in South and South-east Asia. *Environ. Int.* **2009**, *35* (3), 647–654.
- (8) Zhao, F. J.; McGrath, S. P.; Meharg, A. A. Arsenic as a food chain contaminant: mechanisms of plant uptake and metabolism and mitigation strategies. *Annu. Rev. Plant Biol.* **2010**, *61*, 535–59.
- (9) Ng, J. C.; Wang, J.; Shraim, A. A global health problem caused by arsenic from natural sources. *Chemosphere* **2003**, *52* (9), 1353–1359.
- (10) Smith, A. H.; Hopenhayn-Rich, C.; Bates, M. N.; Goeden, H. M.; Hertz-Picciotto, I.; Duggan, H. M.; Wood, R.; Kosnett, M. J.; Smith, M. T. Cancer risks from arsenic in drinking water. *Environ. Health Perspect.* **1992**, *97*, 259.
- (11) States, J. C.; Srivastava, S.; Chen, Y.; Barchowsky, A. Arsenic and cardiovascular disease. *Toxicol. Sci.* **2009**, *107* (2), 312–23.
- (12) Lai, M. S.; Hsueh, Y. M.; Chen, C. J.; Shyu, M. P.; Chen, S. Y.; Kuo, T. L.; Wu, M. M.; Tai, T. Y. Ingested inorganic arsenic and prevalence of diabetes mellitus. *Am. J. Epidemiol.* **1994**, *139* (5), 484–492.
- (13) Tseng, C.; Blackfoot, H. Blackfoot disease and arsenic: a never-ending story. *J. Environ. Sci. Health. C. Environ. Carcinog. Ecotoxicol Rev.* **2005**, *23* (1), 55–74.
- (14) Cullen, W. R.; Reimer, K. J. Arsenic speciation in the environment. *Chem. Rev.* **1989**, *89* (4), 713–764.
- (15) Akter, K. F.; Owens, G.; Davey, D. E.; Naidu, R. Arsenic Speciation and Toxicity in Biological Systems. *Rev. Environ. Contam. Toxicol.* **2005**, *184*, 97–149.
- (16) Vega, L.; Styblo, M.; Patterson, R.; Cullen, W.; Wang, C.; Germolec, D. Differential effects of trivalent and pentavalent arsenicals on cell proliferation and cytokine secretion in normal

human epidermal keratinocytes. *Toxicol. Appl. Pharmacol.* **2001**, *172* (3), 225–232.

(17) Sforza, M. C.; Philippot, P.; Somogyi, A.; van Zuilen, M. A.; Medjoubi, K.; Schoepp-Cothenet, B.; Nitschke, W.; Visscher, P. T. Evidence for arsenic metabolism and cycling by microorganisms 2.7 billion years ago. *Nat. Geosci.* **2014**, *7* (11), 811–815.

(18) Zhu, Y. G.; Xue, X. M.; Kappler, A.; Rosen, B. P.; Meharg, A. A. Linking genes to microbial biogeochemical cycling: lessons from arsenic. *Environ. Sci. Technol.* **2017**, *51* (13), 7326–7339.

(19) Turpeinen, R.; Pansar-Kallio, M.; Kairesalo, T. Role of microbes in controlling the speciation of arsenic and production of arsines in contaminated soils. *Sci. Total Environ.* **2002**, *285* (1), 133–145.

(20) Oremland, R. S.; Stolz, J. F. Arsenic, microbes and contaminated aquifers. *Trends Microbiol.* **2005**, *13* (2), 45–9.

(21) Qin, J.; Rosen, B. P.; Zhang, Y.; Wang, G.; Franke, S.; Rensing, C. Arsenic detoxification and evolution of trimethylarsine gas by a microbial arsenite S-adenosylmethionine methyltransferase. *Proc. Natl. Acad. Sci. U. S. A.* **2006**, *103* (7), 2075–80.

(22) Yan, Y.; Ye, J.; Xue, X. M.; Zhu, Y. G. Arsenic Demethylation by a C.As Lyase in Cyanobacterium *Nostoc* sp. PCC 7120. *Environ. Sci. Technol.* **2015**, *49* (24), 14350–8.

(23) Yang, H. C.; Rosen, B. P. New mechanisms of bacterial arsenic resistance. *Biomed. J.* **2016**, *39* (1), 5–13.

(24) Yoshinaga, M.; Cai, Y.; Rosen, B. P. Demethylation of methylarsonic acid by a microbial community. *Environ. Microbiol.* **2011**, *13* (5), 1205–15.

(25) Xiao, K. Q.; Li, L. G.; Ma, L. P.; Zhang, S. Y.; Bao, P.; Zhang, T.; Zhu, Y. G. Metagenomic analysis revealed highly diverse microbial arsenic metabolism genes in paddy soils with low-arsenic contents. *Environ. Pollut.* **2016**, *211*, 1–8.

(26) Rascovan, N.; Maldonado, J.; Vazquez, M. P.; Eugenia Farias, M. Metagenomic study of red biofilms from Diamante Lake reveals ancient arsenic bioenergetics in haloarchaea. *ISME J.* **2016**, *10* (2), 299–309.

(27) Cai, L.; Yu, K.; Yang, Y.; Chen, B. W.; Li, X. D.; Zhang, T. Metagenomic exploration reveals high levels of microbial arsenic metabolism genes in activated sludge and coastal sediments. *Appl. Microbiol. Biotechnol.* **2013**, *97* (21), 9579–9588.

(28) Zhang, S. Y.; Zhao, F. J.; Sun, G. X.; Su, J. Q.; Yang, X. R.; Li, H.; Zhu, Y. G. Diversity and abundance of arsenic biotransformation genes in paddy soils from southern China. *Environ. Sci. Technol.* **2015**, *49* (7), 4138–46.

(29) Giloteaux, L.; Holmes, D. E.; Williams, K. H.; Wrighton, K. C.; Wilkins, M. J.; Montgomery, A. P.; Smith, J. A.; Orellana, R.; Thompson, C. A.; Roper, T. J.; et al. Characterization and transcription of arsenic respiration and resistance genes during in situ uranium bioremediation. *ISME J.* **2013**, *7* (2), 370.

(30) Zhang, S. Y.; Su, J. Q.; Sun, G. X.; Yang, Y.; Zhao, Y.; Ding, J.; Chen, Y. S.; Shen, Y.; Zhu, G.; Rensing, C.; Zhu, Y. G. Land scale biogeography of arsenic biotransformation genes in estuarine wetland. *Environ. Microbiol.* **2017**, *19* (6), 2468–2482.

(31) Achour, A. R.; Bauda, P.; Billard, P. Diversity of arsenite transporter genes from arsenic-resistant soil bacteria. *Res. Microbiol.* **2007**, *158* (2), 128–37.

(32) Giloteaux, L.; Holmes, D. E.; Williams, K. H.; Wrighton, K. C.; Wilkins, M. J.; Montgomery, A. P.; Smith, J. A.; Orellana, R.; Thompson, C. A.; Roper, T. J.; Long, P. E.; Lovley, D. R. Characterization and transcription of arsenic respiration and resistance genes during in situ uranium bioremediation. *ISME J.* **2013**, *7* (2), 370–83.

(33) Jia, Y.; Huang, H.; Zhong, M.; Wang, F. H.; Zhang, L. M.; Zhu, Y. G. Microbial arsenic methylation in soil and rice rhizosphere. *Environ. Sci. Technol.* **2013**, *47* (7), 3141–8.

(34) Reid, M. C.; Maillard, J.; Bagnoud, A.; Falquet, L.; Le Vo, P.; Bernier-Latmani, R. Arsenic Methylation Dynamics in a Rice Paddy Soil Anaerobic Enrichment Culture. *Environ. Sci. Technol.* **2017**, *51* (18), 10546–10554.

- (35) Cai, L.; Liu, G.; Rensing, C.; Wang, G. Genes involved in arsenic transformation and resistance associated with different levels of arsenic-contaminated soils. *BMC Microbiol.* **2009**, *9*, 4.
- (36) Saltikov, C. W.; Newman, D. K. Genetic identification of a respiratory arsenate reductase. *Proc. Natl. Acad. Sci. U. S. A.* **2003**, *100* (19), 10983–8.
- (37) Sun, Y.; Polishchuk, E. A.; Radoja, U.; Cullen, W. R. Identification and quantification of arsC genes in environmental samples by using real-time PCR. *J. Microbiol. Methods* **2004**, *58* (3), 335–49.
- (38) Whetzel, P. L.; Parkinson, H.; Causton, H. C.; Fan, L.; Fostel, J.; Fragoso, G.; Game, L.; Heiskanen, M.; Morrison, N.; Rocca-Serra, P.; et al. The MGED Ontology: a resource for semantics-based description of microarray experiments. *Bioinformatics* **2006**, *22* (7), 866–873.
- (39) He, Z.; Gentry, T. J.; Schadt, C. W.; Wu, L.; Liebich, J.; Chong, S. C.; Huang, Z.; Wu, W.; Gu, B.; Jardine, P.; Criddle, C.; Zhou, J. GeoChip: a comprehensive microarray for investigating biogeochemical, ecological and environmental processes. *ISME J.* **2007**, *1* (1), 67–77.
- (40) Dieffenbach, C.; Lowe, T.; Dveksler, G. General concepts for PCR primer design. *Genome Res.* **1993**, *3* (3), S30–S37.
- (41) Innis, M. A.; Gelfand, D. H.; Sninsky, J. J.; White, T. J. *PCR Protocols: A Guide to Methods and Applications*; Academic Press, 2012.
- (42) Zhu, Y. G.; Zhao, Y.; Li, B.; Huang, C. L.; Zhang, S. Y.; Yu, S.; Chen, Y. S.; Zhang, T.; Gillings, M. R.; Su, J. Q. Continental-scale pollution of estuaries with antibiotic resistance genes. *Nat. Microbiol.* **2017**, *2*, 16270.
- (43) Ye, J.; Coulouris, G.; Zaretskaya, I.; Cutcutache, I.; Rozen, S.; Madden, T. L. Primer-BLAST: A tool to design target-specific primers for polymerase chain reaction. *BMC Bioinf.* **2012**, *13* (1), 134.
- (44) Nielsen, S. S.; Kjeldsen, P.; Jakobsen, R. Full scale amendment of a contaminated wood impregnation site with iron water treatment residues. *Front. Environ. Sci. Eng.* **2016**, *10* (4), 3.
- (45) Su, J. Q.; An, X. L.; Li, B.; Chen, Q. L.; Gillings, M. R.; Chen, H.; Zhang, T.; Zhu, Y. G. Metagenomics of urban sewage identifies an extensively shared antibiotic resistome in China. *Microbiome* **2017**, *5* (1), 84.
- (46) Sanger, F.; Nicklen, S.; Coulson, A. R. DNA sequencing with chain-terminating inhibitors. *Proc. Natl. Acad. Sci. U. S. A.* **1977**, *74* (12), 5463–5467.
- (47) Korbie, D. J.; Mattick, J. S. Touchdown PCR for increased specificity and sensitivity in PCR amplification. *Nat. Protoc.* **2008**, *3* (9), 1452–6.
- (48) Tardif, S.; Cipullo, S.; Sø, H. U.; Wragg, J.; Holm, P. E.; Coulon, F.; Brandt, K. K.; Cave, M. Factors governing the solid phase distribution of Cr, Cu and As in contaminated soil after 40 years of ageing. *Sci. Total Environ.* **2019**, *652*, 744–745.
- (49) Pei, R.; Kim, S. C.; Carlson, K. H.; Pruden, A. Effect of river landscape on the sediment concentrations of antibiotics and corresponding antibiotic resistance genes (ARG). *Water Res.* **2006**, *40* (12), 2427–35.
- (50) Looft, T.; Johnson, T. A.; Allen, H. K.; Bayles, D. O.; Alt, D. P.; Stedtfeld, R. D.; Sul, W. J.; Stedtfeld, T. M.; Chai, B.; Cole, J. R.; Hashsham, S. A.; Tiedje, J. M.; Stanton, T. B. In-feed antibiotic effects on the swine intestinal microbiome. *Proc. Natl. Acad. Sci. U. S. A.* **2012**, *109* (5), 1691.
- (51) Deng, W.; Wang, Y.; Liu, Z.; Cheng, H.; Xue, Y. HemI: a toolkit for illustrating heatmaps. *PLoS One* **2014**, *9* (11), No. e111988.
- (52) Brennan, C.; Morrison, T. High throughput, nanoliter quantitative PCR. *Drug Discovery Today: Technol.* **2005**, *2* (3), 247–53.
- (53) Stedtfeld, R. D.; Baushke, S. W.; Tourlousse, D. M.; Miller, S. M.; Stedtfeld, T. M.; Gulari, E.; Tiedje, J. M.; Hashsham, S. A. Development and experimental validation of a predictive threshold cycle equation for quantification of virulence and marker genes by high-throughput nanoliter-volume PCR on the OpenArray platform. *Appl. Environ. Microbiol.* **2008**, *74* (12), 3831–8.
- (54) MacNair, M. R.; Cumbes, Q. J.; Meharg, A. A. The genetics of arsenate tolerance in Yorkshire fog, *Holcus lanatus* L. *Heredity* **1992**, *69* (4), 325.
- (55) Meharg, A. A.; Hartley-Whitaker, J. Arsenic uptake and metabolism in arsenic resistant and nonresistant plant species. *New Phytol.* **2002**, *154* (1), 29–43.
- (56) Frick, H.; Tardif, S.; Kandeler, E.; Holm, P. E.; Brandt, K. K. Assessment of biochar and zero-valent iron for in-situ remediation of chromated copper arsenate contaminated soil. *Sci. Total Environ.* **2018**, *655*, 414–22.
- (57) Ye, J.; Rensing, C.; Rosen, B. P.; Zhu, Y. G. Arsenic biomethylation by photosynthetic organisms. *Trends Plant Sci.* **2012**, *17* (3), 155–62.
- (58) Muller, D.; Lievreumont, D.; Simeonova, D. D.; Hubert, J. C.; Lett, M. C. Arsenite oxidase aox genes from a metal-resistant β -proteobacterium. *J. Bacteriol.* **2003**, *185* (1), 135–141.
- (59) Dahl, A.; Sultan, M.; Jung, A.; Schwartz, R.; Lange, M.; Steinwand, M.; Livak, K. J.; Lehrach, H.; Nyarsik, L. Quantitative PCR based expression analysis on a nanoliter scale using polymer nano-well chips. *Biomed. Microdevices* **2007**, *9* (3), 307–14.
- (60) Stedtfeld, R. D.; Williams, M. R.; Fakher, U.; Johnson, T. A.; Stedtfeld, T. M.; Wang, F.; Khalife, W. T.; Hughes, M.; Etchebarne, B. E.; Tiedje, J. M.; Hashsham, S. A. Antimicrobial resistance dashboard application for mapping environmental occurrence and resistant pathogens. *FEMS Microbiol. Ecol.* **2016**, *92* (3), fiv020.
- (61) Zhu, Y. G.; Johnson, T. A.; Su, J. Q.; Qiao, M.; Guo, G. X.; Stedtfeld, R. D.; Hashsham, S. A.; Tiedje, J. M. Diverse and abundant antibiotic resistance genes in Chinese swine farms. *Proc. Natl. Acad. Sci. U. S. A.* **2013**, *110* (9), 3435–40.
- (62) Su, J. Q.; Wei, B.; Ou-Yang, W. Y.; Huang, F. Y.; Zhao, Y.; Xu, H. J.; Zhu, Y. G. Antibiotic resistome and its association with bacterial communities during sewage sludge composting. *Environ. Sci. Technol.* **2015**, *49* (12), 7356–63.
- (63) Morrison, T.; Hurley, J.; Garcia, J.; Yoder, K.; Katz, A.; Roberts, D.; Cho, J.; Kanigan, T.; Ilyin, S. E.; Horowitz, D.; Dixon, J. M.; Brennan, C. J. Nanoliter high throughput quantitative PCR. *Nucleic Acids Res.* **2006**, *34* (18), e123.
- (64) Pfaffl, M. W. Quantification strategies in real-time PCR. In *A-Z of Quantitative PCR*; Bustin, S. A., Eds.; International University Line: La Jolla, CA, USA, 2004; pp 87–112.
- (65) Livak, K. J.; Schmittgen, T. D. Analysis of relative gene expression data using real-time quantitative PCR and the $2^{-\Delta\Delta C(T)}$ Method. *Methods* **2001**, *25* (4), 402–8.
- (66) Schmittgen, T. D.; Livak, K. J. Analyzing real-time PCR data by the comparative CT method. *Nat. Protoc.* **2008**, *3* (6), 1101–1108.
- (67) Lomax, C.; Liu, W. J.; Wu, L.; Xue, K.; Xiong, J.; Zhou, J.; McGrath, S. P.; Meharg, A. A.; Miller, A. J.; Zhao, F. J. Methylated arsenic species in plants originate from soil microorganisms. *New Phytol.* **2012**, *193* (3), 665–72.
- (68) Wang, L.; Chen, S.; Xiao, X.; Huang, X.; You, D.; Zhou, X.; Deng, Z. arsRBOCT arsenic resistance system encoded by linear plasmid pHZ227 in *Streptomyces* sp. strain FR-008. *Appl. Environ. Microbiol.* **2006**, *72* (5), 3738–42.
- (69) Chauhan, N. S.; Ranjan, R.; Purohit, H. J.; Kalia, V. C.; Sharma, R. Identification of genes conferring arsenic resistance to *Escherichia coli* from an effluent treatment plant sludge metagenomic library. *FEMS Microbiol. Ecol.* **2009**, *67* (1), 130–9.
- (70) Chen, J.; Yoshinaga, M.; Garbinski, L. D.; Rosen, B. P. Synergistic interaction of glyceraldehydes-3-phosphate dehydrogenase and ArsJ, a novel organoarsenical efflux permease, confers arsenate resistance. *Mol. Microbiol.* **2016**, *100* (6), 945–53.
- (71) Chrysostomou, C.; Quandt, E. M.; Marshall, N. M.; Stone, E.; Georgioug, G. An alternate pathway of arsenate resistance in *E. coli* mediated by the glutathione S-transferase GstB. *ACS Chem. Biol.* **2015**, *10* (3), 875–82.
- (72) Zhu, Y. G.; Gillings, M.; Simonet, P.; Stekel, D.; Banwart, S.; Penuelas, J. Human dissemination of genes and microorganisms in Earth's Critical Zone. *Glob. Chang. Biol.* **2018**, *24* (4), 1488–1499.

(73) Sardiwal, S.; Santini, J. M.; Osborne, T. H.; Djordjevic, S. Characterization of a two-component signal transduction system that controls arsenite oxidation in the chemolithoautotroph NT-26. *FEMS Microbiol. Lett.* **2010**, *313* (1), 20–8.

(74) Heinrich-Salmeron, A.; Cordi, A.; Brochier-Armanet, C.; Halter, D.; Pagnout, C.; Abbaszadeh-fard, E.; Montaut, D.; Seby, F.; Bertin, P. N.; Bauda, P.; Arsene-Ploetze, F. Unsuspected diversity of arsenite-oxidizing bacteria as revealed by widespread distribution of the *aoxB* gene in prokaryotes. *Appl. Environ. Microbiol.* **2011**, *77* (13), 4685–92.

(75) Zargar, K.; Conrad, A.; Bernick, D. L.; Lowe, T. M.; Stolc, V.; Hoeft, S.; Oremland, R. S.; Stolz, J.; Saltikov, C. W. ArxA, a new clade of arsenite oxidase within the DMSO reductase family of molybdenum oxidoreductases. *Environ. Microbiol.* **2012**, *14* (7), 1635–45.

(76) Malasarn, D.; Saltikov, C. W.; Campbell, K. M.; Santini, J. M.; Hering, J. G.; Newman, D. K. *arrA* is a reliable marker for As(V) respiration. *Science* **2004**, *306* (5695), 455.

(77) Rosen, B. P. Biochemistry of arsenic detoxification. *FEBS Lett.* **2002**, *529* (1), 86–92.

(78) Lin, Y. F.; Walmsley, A. R.; Rosen, B. P. An arsenic metallochaperone for an arsenic detoxification pump. *Proc. Natl. Acad. Sci. U. S. A.* **2006**, *103* (42), 15617–22.

(79) Tisa, L. S.; Rosen, B. P. Molecular characterization of an anion pump. The ArsB protein is the membrane anchor for the ArsA protein. *J. Biol. Chem.* **1990**, *265* (1), 190–194.

(80) Chen, J.; Madegowda, M.; Bhattacharjee, H.; Rosen, B. P. ArsP: a methylarsenite efflux permease. *Mol. Microbiol.* **2015**, *98* (4), 625–35.

(81) Bobrowicz, P.; Wysocki, R.; Owsianik, G.; Goffeau, A.; Ulaszewski, S. Isolation of three contiguous genes, ACR1, ACR2 and ACR3, involved in resistance to arsenic compounds in the yeast *Saccharomyces cerevisiae*. *Yeast* **1997**, *13* (9), 819–828.

(82) Ben Fekih, I.; Zhang, C.; Li, Y. P.; Zhao, Y.; Alwathnani, H. A.; Saquib, Q.; Rensing, C.; Cervantes, C. Distribution of arsenic resistance genes in prokaryotes. *Front. Microbiol.* **2018**, *9*, 2473.