

CHEMICAL ATTRIBUTES OF A DEGRADED SOIL AFTER APPLICATION OF WATER TREATMENT SLUDGES

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ABSTRACT: Mining is an activity that tends to degrade the environment. The restoration of mining areas, aims to accelerate and improve the succession process according to its future use. The objective of this work was to rehabilitate a soil degraded by tin mining activities in the Amazon Basin (Jamari National Forest, State of Rondônia) with the application of water treatment sludge (WTS), and verify the effect of sludge on values of pH (CaCl_2 0.01 mol L⁻¹), organic matter, P, Ca, Mg, K, H+Al, and soil micronutrient contents when cultivated with native plants, legumes, and grass species. A factorial (3 × 5) experimental design was used to optimize the rehabilitation of these areas including three N rates (100, 200, and 300 mg N kg⁻¹ soil supplied by WTS), five plant species (grasses, legumes, and native plants), and two controls (degraded soil with no fertilizer and degraded soil fertilized with mineral fertilizers), with four replications. WTS increased pH values. The chemical products used to treat the water contributed, in greater extension, to increase soil Ca and Fe contents. The use of WTS as fertilizer proved viable, since it contains nutrients for plants; however, nitrogen should not be used as a criterion to define the rate of sludge application, because it is present at small amounts in the WTS.

Key words: soil fertility, micronutrient, tin mining, soil reclaiming, residue

ATRIBUTOS QUÍMICOS EM SOLO DEGRADADO APÓS APLICAÇÃO DE LODO DE ESTAÇÃO DE TRATAMENTO DE ÁGUA

RESUMO: A mineração é uma atividade que tende a degradar o ambiente. A recuperação destas áreas visa acelerar o processo de sucessão e, se possível, melhorá-las de acordo com o uso futuro. O objetivo deste trabalho foi avaliar o efeito da aplicação de lodo de Estação de Tratamento de Água (ETA) em solo degradado pela mineração de cassiterita na Floresta Nacional do Jamari, RO, sobre atributos de fertilidade do solo e sobre o conteúdo em micronutrientes, quando cultivado com plantas pioneiras, leguminosas e gramíneas. O delineamento experimental foi inteiramente casualizado em esquema fatorial 3 × 5 com 4 repetições. Os tratamentos incluíram três doses de nitrogênio (100, 200 e 300 mg de N kg⁻¹ solo), fornecidas pelo lodo de ETA, cinco tipos de plantas (gramínea, leguminosa e pioneira) e duas testemunhas (solo degradado sem e com fertilização mineral). O lodo de ETA aumentou o pH do solo degradado. Os produtos utilizados no tratamento químico da água contribuíram para aumentar, em maior proporção, os teores de Ca e Fe no solo. O uso do lodo de ETA como fertilizante mostrou-se viável, porque contém nutrientes de plantas, no entanto o nitrogênio não deve ser utilizado como critério para definir as doses de aplicação, porque este se encontra em pequena quantidade no lodo de ETA.

Palavras-chave: fertilidade do solo, micronutrientes, mineração de estanho, recuperação, resíduos

INTRODUCTION

Mining is an important activity for regional development. However, because of its extractivist nature, it tends to degrade the environment. The most important loss is observed in the soil's fertile layer, where organic matter, soil fauna, and nutrients are concentrated (Griffith et al., 1990).

The recovery of degraded areas by men, aims to accelerate the succession process and, if possible, improve such areas in accordance with their future use. Grasses

and legumes have been suggested among the species used to begin revegetation (Barth, 1986). Among grasses, *Brachiaria decumbens* has been the most popular because of its ruggedness (Fonseca, 1987; Rocha, 1986). Legumes, such as velvet bean, sunn hemp, and lablab offer great soil protection, respond favorably to the application of fertilizers, and fix atmospheric N. As a mineral fertilizer, N has been an essential element for the development of grasses, including *B. decumbens*, when grown in degraded areas (Silva et al., 1995, Coppin & Bradshaw, 1982). P and K have also played important roles.

Water Treatment Sludge (WTS) is an inorganic residue, containing on average 3% C and 0.6% organic-N (Elliott et al., 1990), so that it could replace, at least in part, the mineral fertilization used in the recovery of degraded areas. This use would help solve another huge environmental problem, the destination of WTS, since the estimated production of this residue in the São Paulo Metropolitan region is about 2,200 L s⁻¹ (SABESP, 1987) and at present all of it is returned to water streams.

The objective of this work was to evaluate the effects of WTS doses applied to a soil degraded by tin mining, and cultivated with a pioneer species associated to grasses and legumes, on soil chemical attributes.

MATERIAL AND METHODS

A sample of Typic Hapludox, collected on the surface layer (0-20 cm) of an area degraded by tin mining at Floresta Nacional do Jamari (Jamary National Forest) (09°15'45" S, 62°49'32" W), RO, Brazil, was air-dried in the shade, sifted (5 mm), and used to fill pots (5 kg capacity). The soil chemical and granulometric attributes (Raij et al., 1997) before the experiment was installed were: pH (CaCl₂ 0.01 mol L⁻¹) = 4.9; OM = 3 g dm⁻³; P(resin) = 8 mg dm⁻³; K = 0.5 mmol_c dm⁻³; Ca = 5 mmol_c dm⁻³; Mg = 2 mmol_c dm⁻³; H+Al = 12 mmol_c dm⁻³; CEC = 19.5 mmol_c dm⁻³; bases saturation = 38%; clay = 170 g kg⁻¹; silt = 30 g kg⁻¹, and sand = 800 g kg⁻¹.

The WTS was obtained from the Water Treatment Station of Araraquara, SP, which uses ferric chloride as coagulant and lime to control pH in the reaction medium. The residue had the following chemical attributes: Organic-C = 10.5 g kg⁻¹, N = 2.0 g kg⁻¹, P = 1.0 g kg⁻¹, K = 2.0 g kg⁻¹, Ca = 121 g kg⁻¹, Mg = 4.0 g kg⁻¹, S = 4.0 g kg⁻¹; Cu = 149 mg kg⁻¹, Fe = 167,040 mg kg⁻¹, Mn = 1,683 mg kg⁻¹, Zn = 66 mg kg⁻¹, B = 5,477 mg kg⁻¹, Cd = 6 mg kg⁻¹, Cr = 86 mg kg⁻¹, Ni = 26 mg kg⁻¹, and Pb = 8 mg kg⁻¹ (dry basis). The granulometric attributes of the WTS were: clay = 425 g kg⁻¹, silt = 315 g kg⁻¹, and sand = 260 g kg⁻¹. Organic C was determined by the wet digestion (Dabin, 1976). Nitrogen was determined in sulfuric digestion extract by steam distillation method (Sarruge & Haag, 1974), while the other nutrients were determined in nitric-perchloric acid digestion extract (except B, obtained by dry digestion) by colorimetry (P), flame photometry (K), atomic absorption spectrophotometry (Ca and Mg), and turbidimetry (S) according to Bataglia et al. (1983). Heavy metals were determined in the extract from digestion with HNO₃, HCl, and H₂O₂ (USEPA, 1986) by atomic absorption spectrophotometry.

A completely randomized experimental design was set up, in a 3 × 5 factorial arrangement, including three nitrogen doses supplied by WTS (D₁₀₀ = 100, D₂₀₀ = 200, and D₃₀₀ = 300 mg N kg⁻¹ soil, dry basis), five types of plants (brachiaria, Tanzania grass, velvet bean, sunn

hemp, and false sicklepod) and two controls (T_s = soil without treatment and T_m = soil + liming + mineral fertilization), with four replicates (68 plots).

For the WTS application, the pots (10 kg capacity, containing 8 kg of degraded soil sieved to 5 mm) received, each other day, 360 g wet sludge in the D₁₀₀ and D₂₀₀ treatments, and 640 g in the D₃₀₀ treatment until the N dose established was reached. On the day following each application, the sludge was superficially incorporated. Considering a 6% solids content and a 2 g kg⁻¹ N content in the sludge on average (dry basis), applications consisted of 5.7, 11.4, and 17.1 kg wet sludge per pot in the D₁₀₀, D₂₀₀, and D₃₀₀ treatments, respectively. After the WTS application was finished, the soil contained in each pot was removed, air-dried in the shade, pounded to break up clods, passed through a sieve (5 mm), and homogenized. Later on, dolomitic lime was applied (67% TNP) to all pots, except for the T_s control, at a rate equivalent to 2 t ha⁻¹; the soil was then homogenized and returned to its original pot, and distilled water was added to reach about 70% of the soil's retention capacity; this moisture content was maintained along the incubation period (25 days) by periodically replacing the water losses by evaporation. On the day plants were sown, the soil in the T_m control pots received mineral fertilization consisting of 300 mg N (ammonium sulfate), 150 mg P (triple superphosphate), and 100 mg K (potassium chloride) per kg soil. The pots also received 30 mL of a micronutrient solution (0.96 g boric acid, 1.97 g copper sulfate pentahydrate, 3.08 g manganese sulfate tetrahydrate, 0.09 g sodium molybdate dihydrate, and 7.33 g zinc sulfate per liter) and 10 mL of a Fe-EDTA solution (Melo et al., 1998). In the WTS treatments, NPK was added to complete the content of these nutrients applied to the T_m treatment. The N and K doses were applied half at sowing and half 30 days after sowing.

The pioneer species false sicklepod (*Senna multijuga*) was then planted in all treatments. The legumes *Stizolobium aterrimum* and *Crotalaria juncea*, and the grasses *Brachiaria decumbens* and *Panicum maximum*, were sowed 120 days later in the pots of the respective treatments. The following plant combinations were prepared for each treatment (D₁₀₀, D₂₀₀, and D₃₀₀): false sicklepod + velvet bean, false sicklepod + sunn hemp, false sicklepod + brachiaria, false sicklepod + Tanzania grass, and false sicklepod alone. Sixty days after the legumes and grasses were sown, the plants were cut near the soil surface, the soil was removed from each pot, and roots were separated from the soil. The soil samples were air-dried in the shade, pounded to break up clods, homogenized, sieved (2 mm), and stored in polyethylene bags maintained in a dry chamber until analyser.

The soil samples were analyzed for soil fertility attributes (Raij et al., 1997), and for Cu, Fe, Mn, and Zn extractable by DTPA solution at pH 7.0 (Lindsay & Norwel, 1978) by atomic absorption spectrophotometry.

Results were submitted to analysis of variance and means were compared by the Tukey's test ($P \leq 0.05$), when the F test was significant ($P \leq 0.01$), according to Pimentel-Gomes (1990).

RESULTS AND DISCUSSION

In the T_m control (soil + liming + chemical fertilization) increasing pH, exchangeable bases (K, Ca, and Mg) and P (resin) contents were recorded, and were higher in relation to the absolute control (T_s), as a result of liming and fertilization. Organic matter content was not different in T_m and T_s (Table 1).

The application of WTS increased 1.8, 2.0, and 1.9 pH units in the soil, respectively for D_{100} , D_{200} , and D_{300} , in relation to T_m (Tables 1 and 2). Even though liming was performed in all treatments, except T_s , these increases could be credited to the WTS, since pH was above the values obtained for T_m . The neutralizing action of the WTS occurs because lime (CaO) is used as coadjuvant during the water treatment process.

Organic matter (OM) contents increased with WTS doses (Table 2). Although WTS is considered an inorganic residue (Elliott et al., 1990), the residue used in this experiment showed a total organic carbon content equal to 10.5 g kg⁻¹ sludge (dry basis). This value, in association with the high doses used as a consequence of the N content in the sludge, helped to increase the amount of carbon in the degraded soil, consequently increasing OM content.

When compared with control, the WTS increased significantly P-resin content in the degraded soil; how-

ever, no differences between doses were detected (Tables 1 and 2). If used at doses lower than those ones used in this experiment, WTS may cause soil P fixation, and a deficiency of this nutrient to plants, which leads to a necessity for mineral fertilizer supplementation (Bugbee & Frink, 1985, cited by Cornwel, 1987). However, the P-resin contents obtained in this work fall within very high (D_{100} and D_{200}) and high (D_{300}) ranges for forest species, and within medium ranges for annual species, according to the interpretation table of Raij et al. (1997).

K, Ca, and Mg contents increased with WTS application, and there were no differences in Ca and K contents between the D_{200} and D_{300} doses (Table 2). The contents obtained for these nutrients are classified as high according to IAC's interpretation table (Raij et al., 1997). However, the Ca contents were much higher than those considered appropriate by Malavolta et al. (1974). WTS is very rich in Ca because calcitic lime, which has higher Neutralizing Power (NP) than dolomitic lime is used in the water treatment, and that can cause problem related to Mg plant nutrition.

High pH values and high calcium concentrations are remarkable characteristics of high-lime soils. These conditions widely interfere with the ion adsorption and solubilization processes, which are responsible for the concentration of nutrients in the soil solution (Moreira et al., 2000). P precipitation, with the formation of calcium phosphate, is one of the reasons for the reduction of this element in the soil solution. Even though the soil presented high pH values, P was available at quantities considered adequate for plant nutrition (Table 2).

Table 1 - Fertility attributes in a soil degraded by tin mining submitted or not to liming and mineral fertilization and cultivation with velvet bean, sunn hemp, brachiaria, Tanzania grass, and with the pioneer species false sicklepod in the controls treatment.

Control	pH CaCl ₂	OM	P (resin)	Ca	Mg	K	H+Al	SB	CEC
		g dm ⁻³	mg dm ⁻³	----- mmol _c dm ⁻³ -----					
T_s^1	5.4 b	3 a	6 b	5 b	1.25 b	2 b	10 b	8.4 a	18.7 b ²
T_m	6.0 a	3 a	16 a	15 a	11.5 a	3 a	12 a	29.2 a	41.4 a

¹ T_s = degraded soil without the addition of fertilizers and corrective amendments. T_m = degraded soil that received liming and chemical fertilizers. ²Means followed by the same letters in the column are not different by the Tukey's test ($P \leq 0.05$).

Table 2 - Fertility attributes of a soil degraded by tin mining after the addition of WTS and cultivation with velvet bean, sunn hemp, brachiaria, Tanzania grass, and the pioneer species false sicklepod.

Treatment	pH CaCl ₂	OM	P (resin)	Ca	Mg	K	H+Al	SB	CEC
		g dm ⁻³	mg dm ⁻³	----- mmol _c dm ⁻³ -----					
D_{100}^1	7.8 b	5 c	23 a	267 b	18 c	2 b	6 a	287.6 b	293 b ²
D_{200}	8.0 a	6 b	21 a	331 a	37 b	4 a	6 a	374.5 a	380 a
D_{300}	7.9 ab	7 a	19 a	334 a	43 a	4 a	6 a	381.7 a	388 a

¹ D_{100} , D_{200} , and D_{300} = addition, respectively, of 100, 200, and 300 mg WTS-N kg⁻¹ soil. ²Means followed by the same letters in the column are not different by the Tukey's test ($P \leq 0.05$).

When increasing Ca contents are not accompanied by increasing in Mg contents, the Ca/Mg ratio may change, with detrimental effects to plant nutrition. At the D_{100} , D_{200} , and D_{300} doses, the Ca:Mg ratio was 15:1, 9:1, and 8:1, respectively. These ratios are above the value that is considered adequate (4:1). However, according to Raij et al. (1997), the Ca:Mg ratio is important when one of these nutrients is at low concentrations in the soil; otherwise, values varying from 1 to above 30 are not detrimental to plant growth.

The interaction between WTS doses \times cultivated plants was only significant for K (Table 3). For D_{100} and D_{300} , soil K content did not vary between treatments containing plants, but for D_{200} it was lower in treatments containing legumes. The legumes showed faster initial development than the grasses and the pioneer plant, uptaking more K, with consequent decrease in their concentration in the soil. K does not take part in precipitation or adsorption reactions because of the effect of Ca. However, the high ratios between Ca and K obtained in the D_{100} , D_{200} , and D_{300} treatments, respectively 134:1, 83:1, and 84:1, may result in K deficiency by inhibition of its absorption (Malavolta et al., 1997).

WTS increased cation exchange capacity (CEC) and sum of bases (SB) and decreased potential acidity (H+Al) (Table 2). According to Raij (1991), increased CEC might be related to the presence of 1:1 clay minerals, of Fe, Al, and Mn oxides, and of organic matter itself, whose negative charges depend on pH. The WTS used in this experiment had high amounts of clay in its composition, as well as high amounts of other elements. Soils of the hydrographic watershed from where water is collected for treatment in Araraquara, SP, are derived from basaltic rocks. It is thus fair to infer that part of this clay has negative charges that may improve the physical properties of the degraded soil. However, the CEC value was obtained by the sum of free cations in solution (K^+ , Ca^{2+} , and Mg^{2+}); since there was an

Table 3 - K in a soil degraded by tin mining submitted to WTS doses and cultivation with legumes, grasses and a native specie.

Treatment	K				
	P1	P2	P3	P4	P5 ¹
	----- mmol dm ⁻³ -----				
D_{100} ²	2 Ab	2 Ab	3 Ac	3 Ab	2 Ab ³
D_{200}	3 Bb	3 Bb	5 Aa	5 Aa	4 Aba
D_{300}	4 Aa	5 Aa	4 Ab	4 Aa	4 Aa

¹P1= velvet bean + false sicklepod, P2 = sunn hemp + false sicklepod, P3 = brachiaria + false sicklepod, P4 = Tanzania grass + false sicklepod, P5 – false sicklepod alone. ² D_{100} , D_{200} , and D_{300} = addition, respectively, of 100, 200, and 300 mg WTS-N kg⁻¹ soil. ³Means followed by the same upper case letter in the row, and lower case letter in the column, are not different by the Tukey's test ($P \leq 0.05$).

excess of free cations, mainly Ca^{2+} , because of the type of the sludge and the doses used, it can be assumed that it was overestimated (Raij et al.; 1987, Raij, 1991).

The micronutrient contents in the control and in the treatments receiving WTS are presented in Tables 4 and 5. According to limits listed in Raij et al. (1997), the contents obtained for Cu, Fe, and Zn are considered high, while those obtained for Mn, as well as those obtained for B at the D_{100} dose, are considered medium. However, for the D_{200} and D_{300} doses, the B contents obtained were low, and may limit plant growth (Table 5), including the values obtained in the controls.

The effect of the interaction between doses \times cultivated plants was significant for Cu and Fe (Tables 6 and 7). At the D_{200} dose, the contents of these elements decreased where the pioneer species false sicklepod was grown. However, plants showed little development in all treatments, eliminating the hypothesis of greater Cu and Fe absorption to explain those reductions.

Fe was found at high amounts in DTPA-extractable form, which can be explained by the presence of ferric chloride in the residue which is originated from the water treatment process. The contents obtained were above those considered adequate by Raij et al. (1997). Among the possible implications for plant nutrition, the excess available Fe in the soil may inhibit Cu and Mn

Table 4 - Micronutrients in a soil degraded by tin mining submitted or not to liming and mineral fertilization and cultivation with velvet bean, sunn hemp, brachiaria, Tanzania grass, and with the pioneer species false sicklepod in the controls treatment.

Treatment	B	Cu	Fe	Mn	Zn
T_s ¹	0.04 b	0.1 b	2 a	0.7 a	0.1 b ²
T_m	0.32 a	0.7 a	3 a	0.9 a	2.7 a

¹ T_s = degraded soil without the addition of fertilizers and corrective amendments. T_m = degraded soil that received liming and chemical fertilizers. ²Means followed by the same letters in the column are not different by the Tukey's test ($P \leq 0.05$).

Table 5 - Micronutrients in a soil degraded by tin mining after the addition of WTS and cultivation with velvet bean, sunn hemp, brachiaria, Tanzania grass, and the pioneer species false sicklepod.

Treatment	B	Cu	Fe	Mn	Zn
	----- mg dm ⁻³ -----				
D_{100} ¹	0.22 a	1.8 b	50 c	3.2 b	4 a ²
D_{200}	0.19 b	2.4 a	75 b	4.4 a	2.8 b
D_{300}	0.13 c	2.6 a	84 a	4.5 a	2.5 b

¹ D_{100} , D_{200} , and D_{300} = addition, respectively, of 100, 200, and 300 mg WTS-N kg⁻¹ soil. ²Means followed by the same letters in the column are not different by the Tukey's test ($P \leq 0.05$).

Table 6 - Copper in a soil degraded by tin mining submitted to WTS doses and cultivation with legumes, grasses and a pioneer specie.

Treatment	Cu content				
	P1	P2	P3	P4	P5 ¹
	----- mg dm ⁻³ -----				
D ₁₀₀ ²	1.5 Ab	2.0 Aa	2.0 Ab	1.7 Ab	1.6 Ab ³
D ₂₀₀	3.0 Aa	2.2 ABa	2.8 Aa	2.7 Aa	1.7 Bb
D ₃₀₀	2.5 Aa	2.6 Aa	2.7 Aa	2.5 Aa	2.8 Aa

¹P1= velvet bean + false sicklepod, P2 = sunn hemp + false sicklepod, P3 = brachiaria + false sicklepod, P4 = Tanzania grass + false sicklepod, P5 – false sicklepod alone. ²D₁₀₀, D₂₀₀, and D₃₀₀ = addition, respectively, of 100, 200, and 300 mg WTS-N kg⁻¹ soil. ³Means followed by the same upper case letter in the row, and lower case letter in the column, are not different by the Tukey's test ($P \leq 0.05$).

Table 7 - Fe in a soil degraded by tin mining submitted to WTS doses and cultivation with legumes, grasses and a pioneer specie.

Treatment	Fe content				
	P1	P2	P3	P4	P5 ¹
	----- mg dm ⁻³ -----				
D ₁₀₀ ²	53 Ab	54 Ab	52 Ab	45 Ab	47 Ab ³
D ₂₀₀	80 Aa	74 Aa	83 Aa	80 Aa	58 Bb
D ₃₀₀	89 Aa	81 Aa	84 Aa	84 Aa	83 Aa

¹P1= velvet bean + false sicklepod, P2 = sunn hemp + false sicklepod, P3 = brachiaria + false sicklepod, P4 = Tanzania grass + false sicklepod, P5 – false sicklepod alone. ²D₁₀₀, D₂₀₀, and D₃₀₀ = addition, respectively, of 100, 200, and 300 mg WTS-N kg⁻¹ soil. ³ Means followed by the same upper case letter in the row, and lower case letter in the column, are not different by the Tukey's test ($P \leq 0.05$).

absorption. Such inhibition effect may also be verified for the Ca/Cu ratio, that is, excess Ca in the soil may inhibit Cu absorption by the plants (Malavolta et al., 1997).

REFERENCES

- BARTH, R.C. Reclamation technology for tiling impoundments – part 2 – revegetation. **Mineral & Energy**, v.19, p.1-25, 1986.
- BATAGLIA, O.C.; TEIXEIRA, J.F.; FURLANI, P.R.; FURLANI, M.C.; GALLO, J.R. **Métodos de análise química de planta**. Campinas: Instituto Agrônômico, 1983. 31p. (Circular, 87).
- COMPANHIA DE SANEAMENTO BÁSICO DO ESTADO DE SÃO PAULO – SABESP. Recuperação de águas de lavagens, tratamento e disposição de resíduos sólidos de ETAs da RMSP. **Revista DAE**, v.47, p.216-219, 1987.
- COPPIN, N.J.; BRADSHAW, A.D. **Quarry reclamation**: the establishment of vegetation in quarries and nonmetal mines. London: Mining Journal Books, 1982. 112p.
- CORNWELL, D.A. **Handbook of practice**: water treatment plant waste management. Denver: AWWA, 1987. 459p.
- DABIN, B. **Curso sobre matéria orgânica do solo**. Parte 1. Análise dos compostos húmicos do solo. Piracicaba: CENA, 1976. 115p.
- ELLIOTT, H.A.; DEMPSEY, B.A.; HAMILTON, D.W.; DEWOLFE, J.R. **Land application of water treatment sludges**: impacts and management. Washington: AWWA Research Foundation, American Water Works Association, 1990. 100p.
- FONSECA, D.M. da Níveis críticos de fósforo em amostras de solos para o estabelecimento de *Andropogon gayanus*, *Brachiaria decumbens* e *Hyparrhenia rufa*. Viçosa: UFV, 1987. 146p. (Dissertação – Mestrado).
- GRIFFITH, J.J.; WILLIAMS, D.D.; FRANCO, R.M.; BRETAS, F.S. Funções múltiplas das florestas. Conservação e recuperação do meio ambiente. In: CONGRESSO FLORESTAL BRASILEIRO, 6., Campos de Jordão, 1990. **Trabalhos voluntários**; resumos. Campos do Jordão: SBS; SBEF, 1990. p.69-77.
- LINDSAY, W.L.; NORWEL, W.A. Development of a DTPA soil test for zinc, iron, manganese and copper. **Soil Science Society of America Journal**, v.42, p.421-428, 1978.
- MALAVOLTA, E.; VITTI, G.C.; OLIVEIRA, S.A. **Avaliação do estado nutricional das plantas**: princípios e aplicações. Piracicaba: POTAFOS, 1997. 319p.
- MALAVOLTA, E.; HAAG, H.P.; MELLO, F.A.; BRASIL SOBRINHO, M.C.O. **Nutrição mineral e adubação de plantas cultivadas**. São Paulo: Pioneira, 1974. 727p.
- MELO, W.J.; MELO, G.M.P.; BERTIPAGLIA, L.M.A.; MELO, V.P. **Experimentação sob condições controladas**. Jaboticabal: FUNEP, 1998. 86p.
- MOREIRA, A.; FRANCHINI, J.C.; MORAES, L.A.C.; MALAVOLTA, E. Disponibilidade de nutrientes em vertissolo calcário. **Pesquisa Agropecuária Brasileira**, v.35, p.2107-2113, 2000.
- PIMENTEL-GOMES, F. **Curso de estatística experimental**. Piracicaba: Nobel, 1990. 468p.
- RAIJ, B. van. **Fertilidade do solo e adubação**. Piracicaba: Ceres, 1991. 343p.
- RAIJ, B. van; QUAGGIO, J.A.; CANTARELLA, H.; FERREIRA, M.E.; LOPES, A.S.; BATAGLIA, O.C. **Análise química do solo para fins de fertilidade**. Campinas: Fundação Cargil, 1987. 170p.
- RAIJ, B. van; CANTARELLA, H.; QUAGGIO, J.A.; FURLANI, A.M.C. **Recomendação de adubação e calagem para o Estado de São Paulo**. 2.ed. Campinas: Instituto Agrônômico; Fundação IAC, 1997. 285p. (Boletim Técnico, 100).
- ROCHA, G.L. Perspectivas e problemas de adubação de pastagens no Brasil. In: MATTOS, H.B.; ERNER, J.C.; YAMADA, T.; MALAVOLTA, E. (Ed.) **Calagem e adubação de pastagens**. Piracicaba: Associação Brasileira para Pesquisa da Potassa e do Fosfato, 1986. p.1-29.
- SARRUGE, J.R.; HAAG, H.P. **Análises químicas em plantas**. Piracicaba: ESALQ/USP, 1974. 56p.
- SILVA, K.E.; VALE, A.B.; BARROS, N.F.; GARCIA, R. Utilização de rejeito de mineração de ferro em combinação com solo para revegetação. **Revista Árvore**, v.19, p.405-14, 1995.
- USEPA. **Test method for evaluating solid waste**. Report Number SW-846. Washington: EPA, 1986. (Code of Federal Regulations 40 CFR Part 503).

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