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MODELING OF HYDROGEOLOGIC CONDITIONS AND GROUND WATER
QUALITY AT AN OIL WELL RESERVE PIT, RICHLAND COUNTY, MONTANA

by

Scott M. Payne

B.S., Northland College, 1985

Presented in Partial Fulfillment of the Requirements for the
Degree of Masters of Science

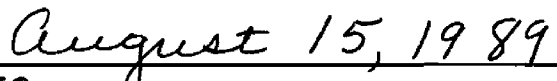
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Geology

MODELING OF HYDROGEOLOGIC CONDITIONS AND GROUNDWATER QUALITY
AT AN OIL WELL RESERVE PIT IN RICHLAND COUNTY, MONTANA

Director: Dr. William W. Woessner *WWW8-7-89*

This study examined groundwater contamination in a shallow aquifer of the Yellowstone River Valley resulting from storage of oil-field brine and drilling fluid additives in a reserve pit. Results show brine seepage from the pit enters the shallow aquifer due to inadequate pit reclamation. Reserve pit mud is contaminated with ions and metals at concentrations one to three orders of magnitude greater than federal drinking water standards. Only certain ions and metals present in the pit mud reach the underlying shallow aquifer system at concentrations much less than in the pit mud.

Chloride concentrations up to 2800 mg/l are present in the shallow aquifer during spring and summer months as a result of the high water table intersecting the base of the reserve pit. In addition, precipitation generates wetting fronts which leach pit contaminants into the shallow aquifer. Chloride seepage declines in the fall through winter months as a result of a lowering water table and less recharge.

After entering the shallow aquifer, chloride migrates downgradient and enters a nearby return flow irrigation ditch. Maximum chloride concentrations measured in the ground water decrease from 2800 mg/l to approximately 600 mg/l at a distance of 350 feet (115 m) from the pit. High chloride concentrations are vertically limited to the upper five to six feet (2 m) of the 15 to 20 foot (6 m) thick sand and gravel aquifer. Lack of vertical mixing in the aquifer is probably a result of a high hydraulic conductivity or layering of strata in the horizontal direction.

Other constituents in the shallow aquifer detected at above background concentrations and attributed to brine seepage from the reserve pit are boron, lithium, barium, strontium, titanium, zinc, beryllium, calcium, magnesium, sodium, potassium, manganese, bicarbonate, sulfate, and nitrates.

Surface electromagnetic induction conductivity (EM) and resistivity were used to delineate the extent of brine contamination. Both methods produced similar results and delineated the brine contaminated area. EM was less cumbersome and time consuming to use compared to resistivity.

Transient groundwater flow modeling for one year of hydraulic head data using the PLASM 2-D numerical model successfully simulated the actual hydraulic heads. Solute transport modeling with Random-Walk successfully simulated chloride concentrations at some monitoring well locations but not at others. The simplicity of the groundwater flow model in conjunction with Random-Walk limitations yielded inaccurate simulated chloride concentrations.

ACKNOWLEDGMENTS

This project is the result of many individual's time and effort. I am deeply indebted for their assistance, knowledge and information. Residents living in the study area were helpful and cooperative throughout the study. I appreciated their patience and willingness to permit monitor well drilling and water quality sampling on their property.

The following organizations provided technical help and data: The Montana Bureau of Mines and Geology, Richland County, the Montana State University Agricultural Experiment Station in Sidney, the U.S. Department of Agriculture in Sidney, the U.S. Soil Conservation Service in Sidney, the U.S. Geological Survey in Helena and The U.S. Forest Service in Missoula. It is my opinion that the staff members of these public agencies provided outstanding assistance and were accessible and helpful in all of my requests.

A special thanks is due for Randy Skov, my field technician, who braved winter days to collect water quality data and water levels. Joe Donovan of the Montana Bureau of Mines and Geology also deserves a special thanks for his help in the field.

Key individuals in the project were my thesis committee members: Drs. Woessner, Thompson, Potts, and Sonderegger, and project supervisor Jon Reiten of the Montana Bureau of Mines and Geology.

I am deeply indebted to Dr. Sonderegger who proposed and secured the funding for this project and Jon Reiten who directed this project. Jon provided field assistance and hydrogeological knowledge that was instrumental and a necessity to pull this project together. I can't thank him enough. Dr. Woessner provided the modeling and hydrogeologic background from which I have learned from and will benefit from in the future. Dr. Woessner's knowledge has made me a better hydrogeologist and I am grateful. I appreciate the other key individuals' patience and wisdom.

Last but certainly not least, I want to thank Ann who is my best friend and love. Ann's patience, understanding, and support made all the work in this project worthwhile. Without her it would have been difficult.

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Chapter I

INTRODUCTION

Groundwater contamination from an oil well reserve pit was studied in a shallow aquifer of the Yellowstone River Valley, eastern Montana. Significant contamination by chloride and other ions and metals was documented. The source of contamination is a buried reserve pit containing extremely elevated concentrations of the ions and metals. Computer modeling was used to simulating groundwater flow and solute transport of chloride contamination through the shallow aquifer. The following discussion will familiarize the reader with a reserve pit. That discussion is followed by a description of oil production in the Williston basin, goals and objectives of this study, location and description of the study site, and a discussion of previous work done at the study site.

Drilling fluids and brine water are contained in a reserve pit during drilling of oil and gas wells in the oil producing areas of the United States. Pit dimensions vary, but average 150 to 200 feet long (46 to 61 m), 60 to 70 feet wide (18 to 21 m), and 8 to 12 feet deep (2 to 4 m). Most reserve pits are lined with a synthetic or clay liner to protect underlying groundwater from brine seepage, however, in some instances pit liners are not used depending on the geologic setting and location. Contamination of shallow aquifers by reserve pits has been documented in Montana,

Oklahoma, Alabama, North Dakota, Texas, Arkansas, Wyoming, Ohio, and Kansas (Beal, 1986; Beal and others, 1987; Collins, 1971; Dewey, 1984; Fryberger, 1975; Hicks, 1983a and 1983b; Knowels, 1965; Leonard, 1965; McMillion 1965; Murphy and Kehew, 1984; Murphy et. al., 1985, Novak and Yoram, 1988).

Techniques used to reclaim reserve pits usually involve removal of the brine water from the pit (disposed via injection well) and on site disposal of the more viscous drilling fluid or pit mud. On site mud disposal consists of digging trenches radially away from the reserve pit and pushing the pit mud into the trenches. Backhoes used to dig the trenches destroy the integrity of the existing pit liner. After the mud is pushed into the trenches, the site is resurfaced to ground level.

Surface recharge at a reclaimed reserve pit site may generate leachate and contaminate the underlying groundwater. In addition, high groundwater may come in contact with the bottom of the buried reserve pit or trenches and contaminate the groundwater.

Presently, drilling fluid wastes are designated as nonhazardous under the RCRA (Resource Conservation and Recovery Act). According to Kemblowski and Deeley (1987), this exemption is primarily based on the "...low toxicity of drilling fluid wastes". However, reserve pit studies in the last five to six years have shown that oil-field brine (Table 1) and drilling fluid additives (Appendix A) contain products

that are a threat to water, soil and vegetation (Hicks 1983a).

Oil production in the Williston Basin

As of 1982, there were over 4000 oil wells in the Williston Basin (Dewey, 1984). Chloride concentrations in brine and drilling fluids (Murphy and Kehew, 1984) are typically $300,000 \pm 20,000$ mg/l in north central North Dakota and 100,000 to 200,000 mg/l in eastern Montana (Dewey, 1984). Pit reclamation in the Williston Basin most often includes removal of brine water and on site mud disposal in excavated trenches (Dewey, 1984). Contamination of groundwater by brine seepage is reported at all reserve pit sites studied in the Williston Basin (Dewey, 1984; Murphy and Kehew, 1984; Hicks, 1983a and 1983b; and Beal, 1986). Dead or stunted vegetation has been documented at a number of these brine contaminated sites (Hicks, 1983a; Dewey, 1984). Many sites outside of Montana and North Dakota have also experienced dead or stunted vegetation in addition to contaminated drinking water and surface water in streams and rivers (Leonard, 1965; McMillion, 1965; Knowles, 1965).

The processes that leach reserve pit wastes into the shallow groundwater are poorly understood and regional impacts to eastern Montana groundwater are relatively unknown.

Goals and objectives

Goals of this study are: 1) to characterize the mechanism(s) controlling reserve pit waste migration into the shallow aquifer, and if the process(es) is continuous or

TABLE 1. Comparison of dissolved solids in seawater and oil field brine (USEPA, 1972).

Element	Seawater, mg/l.	Oil-field brine, mg/l.
Sodium	10,600	12,000 to 150,000
Potassium	380	30 to 4,000
Lithium	0.2	1 to 50
Rubidium	0.12	0.1 to 7
Cesium	0.0005	0.01 to 3
Calcium	400	1,000 to 120,000
Magnesium	1,300	500 to 25,000
Strontium	8	5 to 5,000
Barium	0.03	0 to 1,000
Chlorine	19,000	20,000 to 250,000
Bromine	65	50 to 5,000
Iodine	0.05	1 to 300

episodic; 2) to determine if EM (electromagnetic induction conductivity) can be used to identify groundwater contaminated from brine seepage; 3) to estimate the aquifer parameters necessary to conduct a hydrogeologic model for the reserve pit area and predict contaminant dispersion behavior; 4) to characterize the chemistry of brine contaminated water and soil and 5) to suggest possible remediation measures. Objectives related to meeting the above goals consisted of measuring/determining the following parameters:

- 1) The soil and groundwater conductance/resistivity using EM and resistivity surface geophysics;
- 2) Groundwater and surface water quality;
- 3) The extent of brine contamination including related metal contamination;
- 4) The hydraulic properties of the unconfined aquifer;
- 5) The rate and direction of groundwater flow at the site;
- 6) The dispersion and dilution of contaminants in the unconfined aquifer;
- 7) Amount of chloride loaded into the groundwater;
- 8) Regional loading of chloride into the Yellowstone River;
- 9) The reclamation technique best suited to reverse soil contamination on site;

Location and description of study site

The Iverson site is located approximately 3 miles north of Sidney, Montana in Section 15 of Township 23 north, Range 59 east in Richland County (Figures 1 and 2). The site is

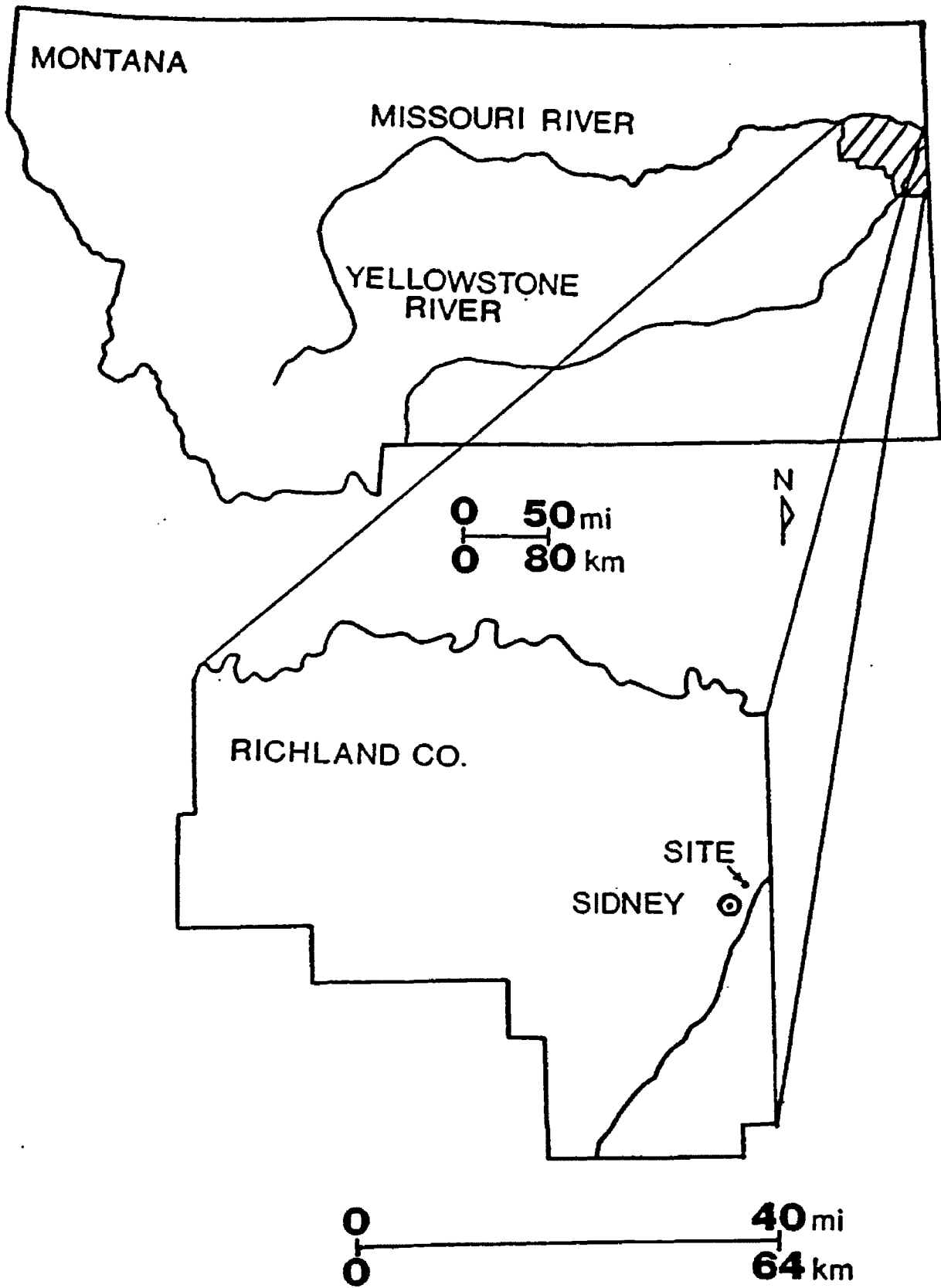


Figure 1. Site delieation in Richland County, Montana.

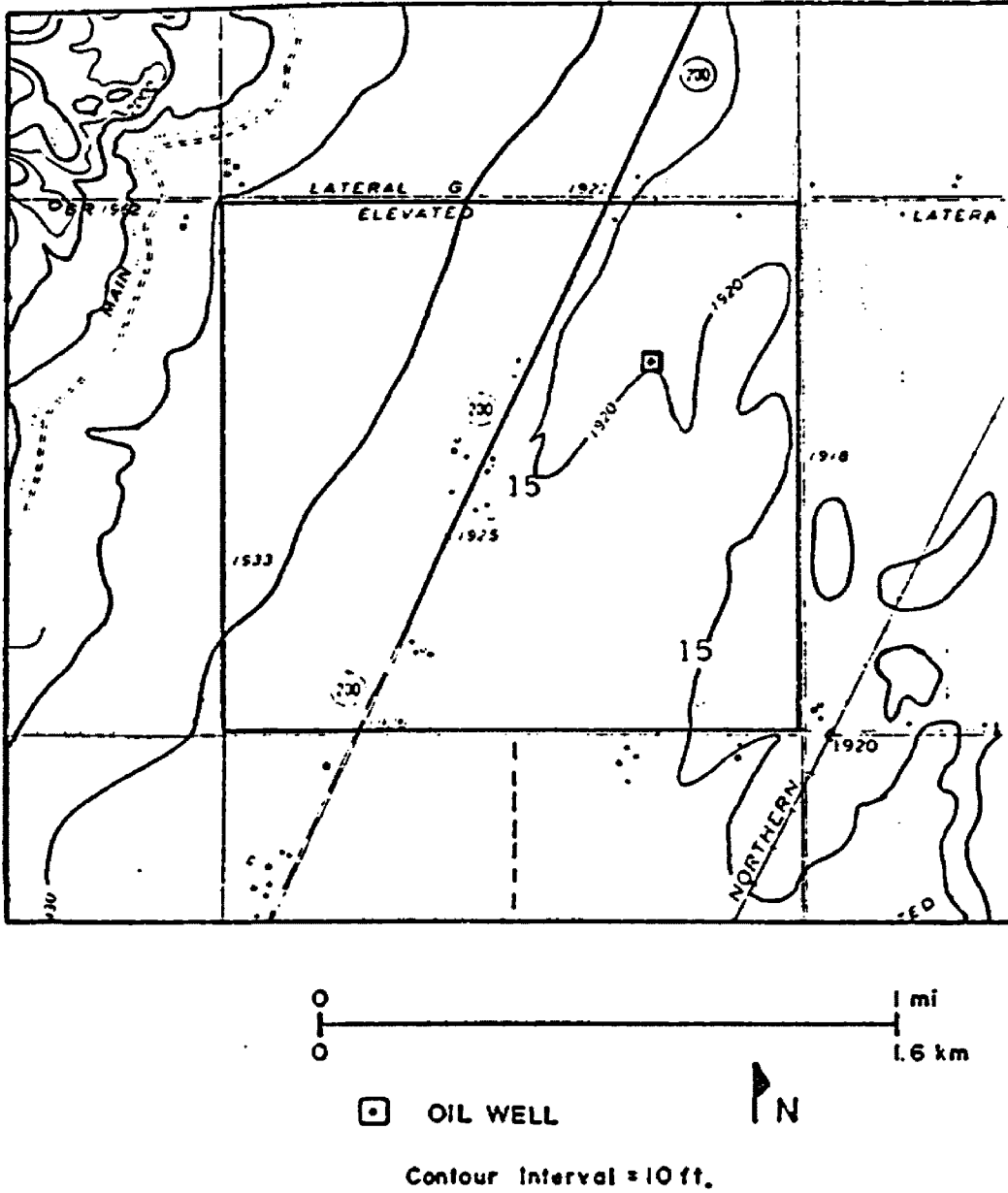


Figure 2. Topography of Section 15, T.23N., R.59E. and location of study site (adapted from Dewey, 1984).

situated on a terrace deposit on the western side of the Yellowstone River ("Crane Creek Gravel"), one mile (1.7 km) east of the western valley edge. The valley edge is composed of the Tongue River member of the Fort Union Formation which is 300 to 500 feet (120 m) higher in elevation than the valley bottom. The Yellowstone River Valley bottom is fairly level, except for a series of low terrace deposits.

At the study site an operating oil pumper and buried reserve pit are present (Figure 3). The reserve pit contained brine water and drilling mud, and was reclaimed in the summer of 1982 using the trenching method (Dewey, 1984). Brine water was removed from the pit and the drilling mud was disposed in both the reserve pit and trenches. According to Dewey (1984), initial groundwater contamination from brine seepage occurred within approximately 20 days after the pit was reclaimed. Wheat, corn, sugar beets, and alfalfa are grown on surrounding land and a return flow irrigation ditch abuts the west side of the oil pumper area. A number of small irrigation channels supply water for flood irrigation near the site.

Climate

The climate is semiarid, characterized by cold, dry winters, moderately hot and dry summers, and cool, dry falls (Slagle, 1984a). Winters are often interrupted by warming trends, with summers dominated by hot days and cool nights. January is generally the coldest month and July the warmest. Glendive, Montana has a 14.9° F average January temperature

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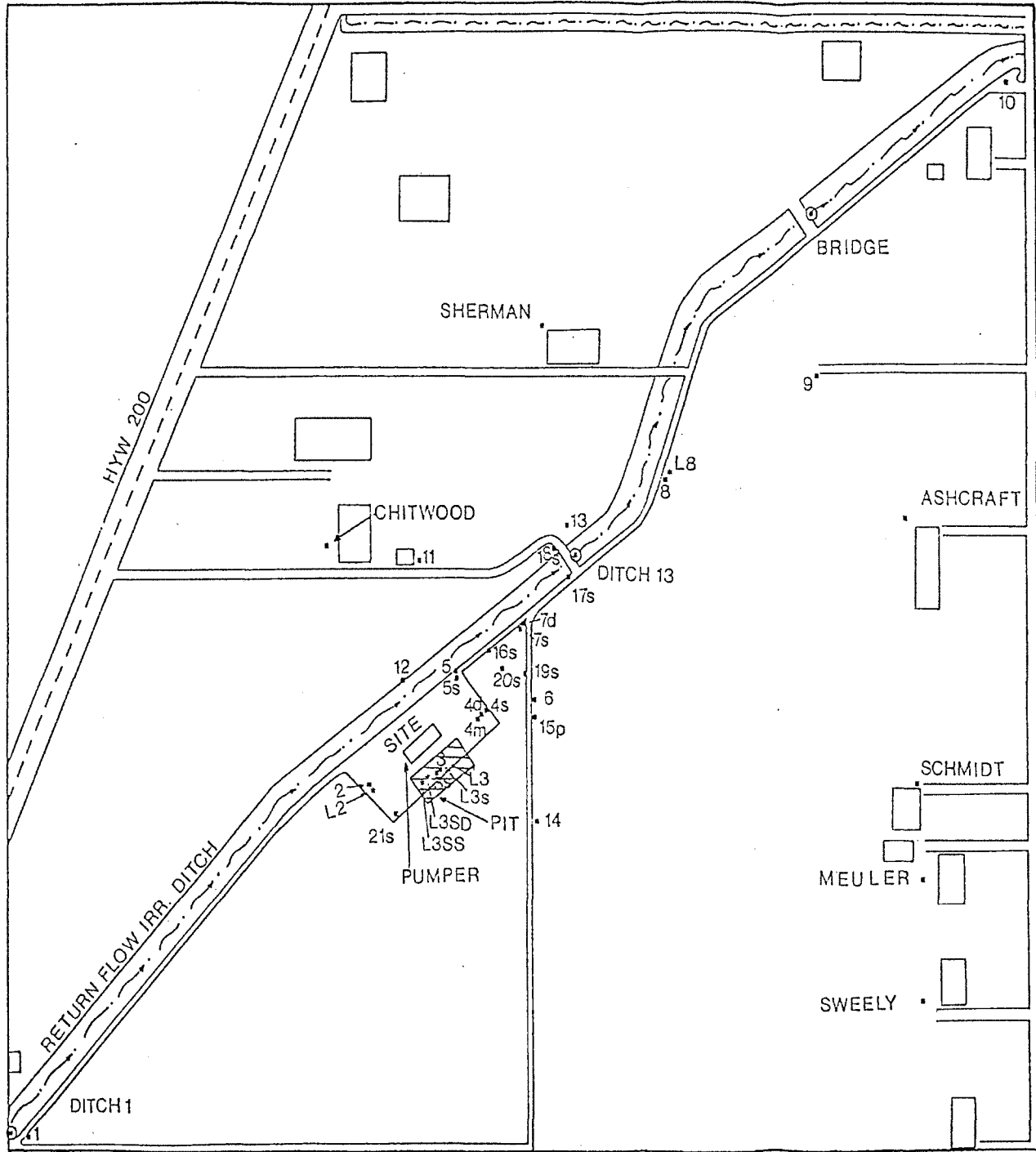
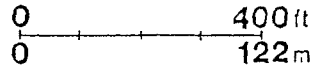


Figure 3.



IVERSON SITE

- WELL
- LYSIMETER
- ⊙ DITCH
- == ROAD
- HOUSE



and a 74.0° F average temperature for July, -10°C and 23.3°C respectively (Slagle, 1984). Average annual precipitation at Sidney is approximately 14 inches/year (35.6 cm) with about 65 percent of the precipitation falling from May through August, and June being the wettest month (Slagle, 1984a).

Regional hydrogeology of Richland County, Montana

Groundwater is present in eight major aquifer systems in Richland County from the Paleozoic Madison Group to the Quaternary alluvium. Aquifers present within 200 to 300 feet (61-91 m) of ground surface in Richland County are the Tongue River Member of the Tertiary Fort Union Formation and the Quaternary glacial, terrace and alluvial deposits. The water table in these aquifers usually reflects the land surface topography (Slagle, 1984a). Deeper aquifers, below 300 feet (91 m), have a regional flow direction towards the Yellowstone and Missouri Rivers.

Groundwater recharge generally occurs in the topographically high areas with discharge areas in the valley bottoms. Snow melt and precipitation are the major contributors of recharge in eastern Montana (Torry and Kohout, 1956), however, irrigation in Yellowstone River Valley contributes large quantities of water to the shallow aquifer(s). Groundwater recharge is greatest in the spring when snow melt and precipitation are at their peak and irrigation begins (Torry and Kohout, 1956).

Prior to this study, no aquifer property data were

available for the Quaternary aquifer systems in Richland County, and only laboratory estimates of hydraulic conductivity were available for the Tongue River Member. Croft (1985) estimated hydraulic conductivities of the Tongue River claystone to be 0.15 to 0.099 ft/day (8.45×10^{-6} to 5.64×10^{-6} cm/sec).

A detailed discussion on the regional hydrogeology is in Appendix B.

Previous work

The Iverson site was previously studied by Michelle Dewey of the University of Montana from August 1982 through June 1983. Dewey determined that a shallow sand and gravel aquifer underlies the site, average depth to the groundwater is nine feet and groundwater flows towards the north.

Dewey's chemical analyses and surface resistivity surveys led her to conclude that: 1) brine is seeping into the shallow groundwater from the reserve pit which produces increased levels of sodium and chloride; 2) the reclamation techniques used to reclaim reserve pits are inadequate to inhibit brine migration; 3) surface electrical resistivity is a good indicator of possible brine contamination at oil well sites; and 4) contaminants may be moving as fast as 22 feet per day through the shallow aquifer.

Numerous other authors have studied reserve pit brine migration in the oil producing areas of the United States. These studies demonstrate that chloride and other ions often

contaminate shallow aquifers and surface water systems comparable to the reserve pit contamination in the shallow aquifer of the Williston Basin (Powell and others, 1963; Knowles, 1965; Fryberger, 1972 and 1975; Todd and McNulty, 1976; Vander Leeden and others, 1975; Latta, 1963; Leonard, 1964; Bryson and others, 1966; Kreiger and Henderickson, 1960; Miller and others, 1977; McMillion, 1965; Scalf and others, 1975; Shaw, 1966; Pettyjohn, 1971, 1973, and 1975; McMillion, 1965; Payne, 1966; Miller; 1980; Baker and Brendecke, 1983).

Chapter II

METHODS

Analytical methods were selected to determine the movement and chemical composition of groundwater and surface water. Monthly soil moisture and groundwater quality sampling methods were selected to determine how and when contaminants enter the groundwater. Computer modeling methods were selected to simulate groundwater flow and solute transport in two dimensions.

Field methods - Geophysical techniques

Resistivity measurement

A Soiltest R-60 resistivity meter and Wenner electrode configuration with electrode a-spacings of 10, 20 and 30 feet (3.48 m, 6.1 m, and 9.57 m, respectively) were used to determine soil and groundwater resistivity. Figure 4 shows the location of resistivity stations. Standard operating procedures as described by Dejong and others (1979) and Soiltest (1976a and 1976b) were used to determine soil and groundwater resistivity.

EM measurement

A Geonics EM34-3 conductivity meter with 10, 20 and 40 metre coil spacings (32.8 ft, 65.6 ft, and 131.23 ft respectively) was used to determine soil and groundwater conductivity in both the horizontal and vertical EM modes. Location of EM stations are shown in Figure 5.

Use of two coil configurations (or modes) with the EM

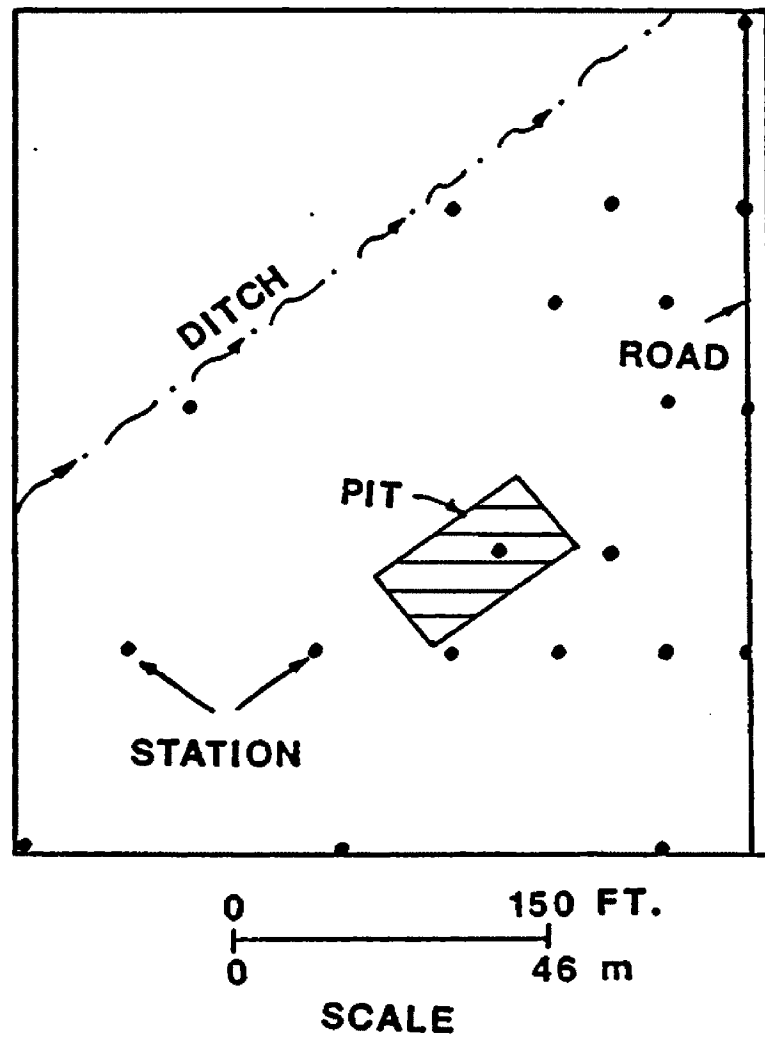
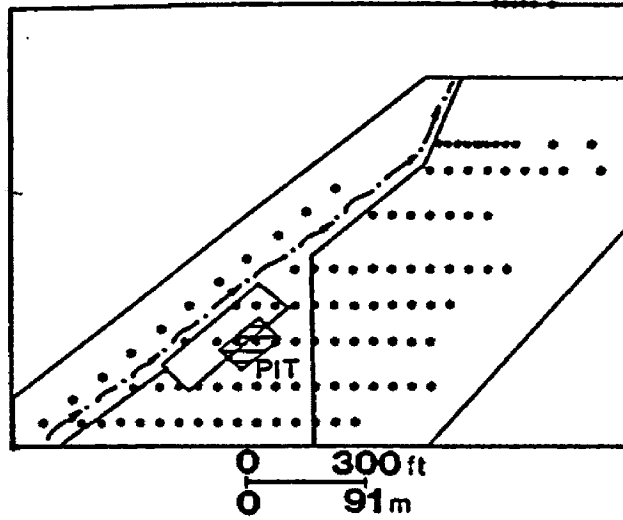
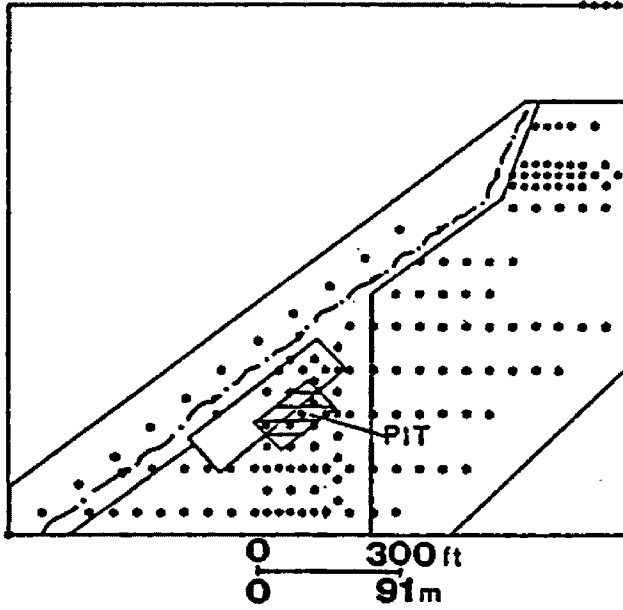


Figure 4. Resistivity grid (10, 20 and 30 ft. a-spacings).

STATION MAP FOR 20 METRE EM



STATION MAP FOR 10 METRE EM



STATION MAP FOR 40 METRE EM

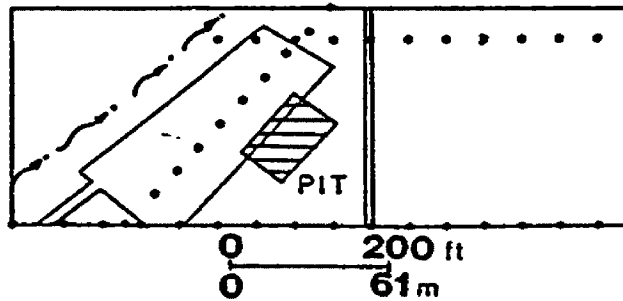


Figure 5. EM gird (10, 20 and 40 metre coil spacings).

equipment measures different depths of soil conductivity. The horizontal EM mode (vertical dipoles) measures soil conductivity to approximately 1.5 times the length of the coil spacing and is relatively insensitive to surface soil conductivity (McNeil, 1980c). The vertical EM mode (horizontal dipoles) measures soil conductivity from the surface to approximately .75 times the length of the coil spacing (McNeil, 1980c).

EM data were interpreted using standard procedures described by McNeil (1980a, 1980b, 1980c, 1985 and 1986).

Monitoring wells and lysimeters:

Figure 3 shows the location of monitoring wells and lysimeters. Appendix C details well and lysimeter construction. The monitoring wells were located on the basis of EM and resistivity results. Readings of large conductivity values and small resistivity values at geophysical survey points were interpreted to represent areas of reserve pit contamination. Monitoring wells were installed using the MBMG (Montana Bureau of Mines and Geology) Mobile-Drill 50 drill rig equipped with 6.5 inch OD (outer-diameter)/3.25 inch ID (inner-diameter) HSA (hollow-stem-auger) flights. One well, MW 15p, was installed using 10 inch OD/6.25 inch ID HSA flights.

The depth of bore holes varied from 10 to 37 feet (3 to 11 m), however, no monitoring wells were installed below 23 feet (7 m). All monitoring wells, except MW 15p, are 2-inch

ID diameter schedule 20 PVC with either a 20 or 30 slot (0.020 or 0.030 inch) well screen. Well MW 15p is a 4.5 inch ID schedule 40 PVC well with a 20 slot (0.020 inch) well screen. The wells are completed with a natural sand and gravel (hole collapse) and a surface bentonite plug. Both single and nested well sites were installed. Nested well sites consisted of two to three monitoring wells completed at different depths a few feet apart. Each well was developed with a hand bailer or pumped with a small surface pump until clear water was obtained. Soil samples and logs were collected from auger cuttings and a geologic log was prepared for each hole (Appendix D).

Monitoring wells installed in the pit area were completed with two bentonite seals. A second bentonite seal (below the surface bentonite) was installed below the bottom of the reserve pit mud to inhibit vertical brine migration along the well casing.

Lysimeters were installed from 3.5 to 6 feet (1 to 2 m) below ground surface. Lysimeters are 1.5 inch diameter schedule 26 PVC with a ceramic cup at the lower end of the casing. The lysimeters are completed with a saturated silica flour surrounding the ceramic cup, natural fill above the silica flour, and a surface bentonite seal.

Water levels were measured with either a steel tape or electric M-scope from a top-of-casing datum surveyed by the United States Soil Conservation Service in Sidney, Montana.

Water levels were measured monthly from July, 1987 to August 1988. Water level data were used to determine head at each well site and areal groundwater flow direction.

A stilling well was installed with a 30 day Stevens continuous water level recorder on the return flow irrigation ditch at site Ditch 1 (Figure 3). Two additional 30 day Stevens continuous recorders were installed on monitoring wells MW 1 and MW 3.

Aquifer testing:

The hydraulic conductivity of the shallow, unconfined sand and gravel aquifer was estimated from an aquifer test conducted at well MW 15p. The shallow aquifer was stressed using a 3 HP surface pump which sustained a constant discharge rate of 30 gpm (gallons/minute) for 4.5 hours. Discharge water was pumped through a pipe approximately 150 feet away to the southwest of the pumping well onto an alfalfa field. Drawdown and recovery were measured in the pumping well (MW 15p) and three monitor wells, 53 feet, 82 feet and 118 feet (16 m, 25 m and 36 m) from the pumping well (wells MW 6, MW 19s, and MW 20s, respectively). Drawdown versus time data were first analyzed by assuming steady-state conditions (Driscoll, 1986). The data were also evaluated assuming transient conditions with delayed yield and partial well penetration (Boulton 1954a, 1954b, 1964, and Stallman, 1965).

Water quality sampling:

Measured field parameters include corrected SC (specific

conductance), temperature, pH, and chloride. Chloride concentrations were measured using Quantabs (Appendix E), a small capillary tube that function as a chloride titrator. A single Quantab is lowered into a cup of water, after which water moves up the capillary passage on the Quantab to discolor the top. A value corresponding to the distance the sample water moved up the capillary passage is read off the side of the Quantab and converted into mg/l or percent chloride in solution (Appendix E). Field data were collected monthly and input into a computer data base.

Water samples for chemical analyses were collected on a quarterly basis. Sampling procedure included bailing water from monitoring wells until pH stabilized and rinsing sample bottles three times with sample water. Four samples were collected from each monitoring well site including one untreated, one filtered, and two filtered and acidified samples. Acidified samples were preserved with 2.5 milliliters of nitric acid per 500 ml of sample water. The filtered samples were pumped through a 0.45-micron filter. Decontamination of sampling equipment was accomplished with three rinses of deionized water.

Analyses for major ions and metals were done at the MBMG laboratory. Metals were also analyzed at the University of Montana Geology Department. Field quality control was checked with one deionized water sample filtered and preserved in the field, and analyzed at the MBMG laboratory. Sample water was

stored in a cooler over ice for delivery to the MBMG laboratory in Butte and the University of Montana Geology Department in Missoula.

Lysimeters were also sampled quarterly. Vacuum was drawn on lysimeters five to 30 days prior to sampling, after which, the soil moisture was sucked out using a vacuum pump. Sample bottles were rinsed with deionized water. The sample water was filtered in the field and sometimes acidified in the field if sufficient sample water was available. Sample bottles were stored in coolers over ice and delivered to the MBMG laboratory and the University of Montana Geology Department.

Laboratory methods - Chemical analyses:

Groundwater and soil moisture samples were analyzed at the MBMG laboratory for following ions and constituents including: calcium, magnesium, sodium, potassium, iron, manganese, silica, bicarbonate, carbonate, chloride, sulfate, nitrate, fluoride, phosphate, bromide, dissolved solids, sum of constituents, specific conductance, pH, hardness, alkalinity, Ryznar stability, Langlier saturation, and sodium absorption ratio. All MBMG laboratory procedures were EPA and USGS certified procedures (Reiten, 1988). Metals analyzed at the MBMG laboratory on an ICPS (inductively coupled plasma spectrometer) were: silver, aluminum, arsenic, boron, beryllium, cadmium, chromium, copper, lithium, molybdenum, barium, lead, nickel, strontium, titanium, vanadium, zinc, zirconium, and selenium.

Metal analyses at the University of Montana were also done on a ICPS. Groundwater, soil moisture, and soil samples were tested for: aluminum, arsenic, calcium, cadmium, copper, iron, potassium, magnesium, manganese, nickel, sodium, lead, phosphate, antimony, and titanium. Soil samples were acid extracted for total digestible metal content (Moore, 1988). Metal scans and sample preparation were accomplished using standard laboratory techniques done in accordance with USGS water analysis standards which included a 10 percent rerun of samples as duplicates (Moore, 1988).

Sieve analysis:

Sieve analyses were performed on three different samples collected from monitoring well cuttings. The samples were sieved at the United States Forest Service Materials Testing Lab in Missoula, Montana. Standard laboratory techniques were used to determine the grain size distribution.

Aquifer porosity was determined by saturating a volume of oven dried aquifer material and measuring the volume of water needed to fill the empty void spaces.

X-ray diffraction:

X-ray diffraction was used to identify the <2-micron size fraction clay mineralogy of mud samples taken from auger cuttings while drilling through the reserve pit. The samples were mixed with deionized water and centrifuged to separate the <2-micron size fraction. The <2-micron size fraction was put on glass slides as random, oriented, and glycol saturated

clay particles for x-ray diffraction (Jennings and Thompson, 1986). Standard X-ray diffraction techniques were used to identify the clay mineralogy of the pit mud.

Computer modeling:

The 2-dimensional PLASM groundwater flow model developed by Prickette and Lonquist (1971) was used to model groundwater flow. Random-Walk, developed by Prickette, Naymik, and Lonquist (1981), was used to model solute transport. PLASM is a finite difference model which can be used to simulate non-steady flow in a heterogeneous aquifer under unconfined and non-leaky, and leaky-confined conditions. PLASM estimates head using the alternating direction implicit numerical method to solve a set of finite difference equations. Input data will be discussed in detail in Chapter III.

Random-Walk is a 2-dimensional model that can simulate contaminant transport in an unconfined or confined aquifer (Prickette and others, 1981). The model takes into account groundwater velocity, convection, dispersion, and chemical reactions. The model simulates movement of a solute using a particles in a cell method, transported for a given time schedule. Particles are randomly dispersed according to model inputs by a statistical technique. The model has the ability to install sinks or sources, and can map solute as concentration or the number of particles for a given cell. The version of Random-Walk used in this study is linked with

the PLASM groundwater flow model. Head data generated with PLASM is converted into a velocity vector field with a preprocessor for each simulated month. The velocity vector field is then linked to the Random-Walk model to move solute in accordance with the modeled potentiometric surface and dispersion coefficients.

Chapter III

RESULTS AND DISCUSSION

This chapter presents all data collected in this study. A short discussion will follow the data in each section.

Site hydrogeology

Soil and fill (material used to cover the pumper area and roads) cover the top three feet (1 m) of land surface of the study area. From three to 26 feet (1 to 8 m) Quaternary sand and gravel of the oldest terrace deposit ("Crane Creek Gravel") of the Yellowstone River Valley is present. Included in the sand and gravel are local zones of well sorted sand, gravel, thin lenses of clay, layered silt and sand, layered clay and sand, and pit mud in the reserve pit area (see geologic cross sections in Appendix D). The depositional environment of the terrace deposit was apparently a fairly high energy system since coarse gravel was deposited with sand. Terrace deposition apparently took place during an interglacial time since till deposits are present stratigraphically above and below the terrace deposits about three miles (5 km) south of the study site (Prichard and Landis, 1975).

A sieve analysis on the sand and gravel material (Appendix D) shows an even distribution of particles, from fine sand to coarse gravel. This material is the dominant aquifer material encountered in the monitoring well bore holes. A sieve analyses on a sample collected from eight to

20 feet (2.5 to 6 m) at monitoring well MW 12 showed that 85% of the sample retained was sand size. A sieve analyses on a sample collected from 14 to 33 feet (4.5 to 10.5 m) at monitoring well MW 13 showed 75% of the sample retained was gravel size. Results of the sieve analyses indicate that the aquifer is relatively coarse grained and little silt and clay are present.

The Tongue River Member of the Fort Union Formation underlies the terrace deposits at approximately 26 to 28 feet (8 m) below ground surface. At the study site the Tongue River Member consists of a dense, low permeability grayish clay with sandy intervals.

The shallow aquifer is restricted to the sand and gravel terrace deposits. The shallow aquifer has a seven to 10 foot (3 m) unsaturated zone and a saturated thickness of about 17 feet (5 m). Below the shallow aquifer, clay of the Tongue River Member separates the shallow unconfined aquifer and the deeper confined aquifer. The upper clay of the Tongue River Member has a relatively low hydraulic conductivity, estimated to be 0.15 to 0.099 ft/day (8.45×10^{-6} to 5.64×10^{-6} cm/sec) (Croft, 1985), which probably inhibits groundwater movement between the shallow and deep aquifers. Domestic wells in the study are completed in water bearing coal and sandy intervals of the Tongue River Member 35 to 45 feet (13 m) below land surface (as determined from domestic well logs).

In general, groundwater flow direction (Figure 6 and

Appendix F) is towards the northeast with some variations in flow direction along the return flow irrigation ditch. The ditch is recharged by groundwater in the spring, summer and fall and the water table is in contact with the ditch since it is above the ditch water level and a visible seepage face is present in the ditch channel. Estimated ditch flow in the summer months is approximately three to four cfs.

In the winter, groundwater flow (Appendix F) is parallel to the ditch, and water in the ditch apparently leaks into the underlying shallow aquifer since the measured ditch water levels are above the water table. The ditch bottom is comprised of mud and silt, and permeability of these sediments is estimated to be relatively low. The low permeability and small winter ditch flow (approximately .5 cfs) indicate that the ditch is most likely leaking water to the shallow aquifer and forming a small groundwater mound.

The Yellowstone Irrigation District shuts down the irrigation canals in the winter months, yet flow is present in the return flow ditch. The source of winter ditch water is probably from upgradient reaches where the water table intersects the ditch causing groundwater to discharge into the ditch.

Precipitation and recharge

Precipitation and free water evaporation data collected at the Montana State University Agricultural Experiment Station in Sidney about two miles from the study area are in

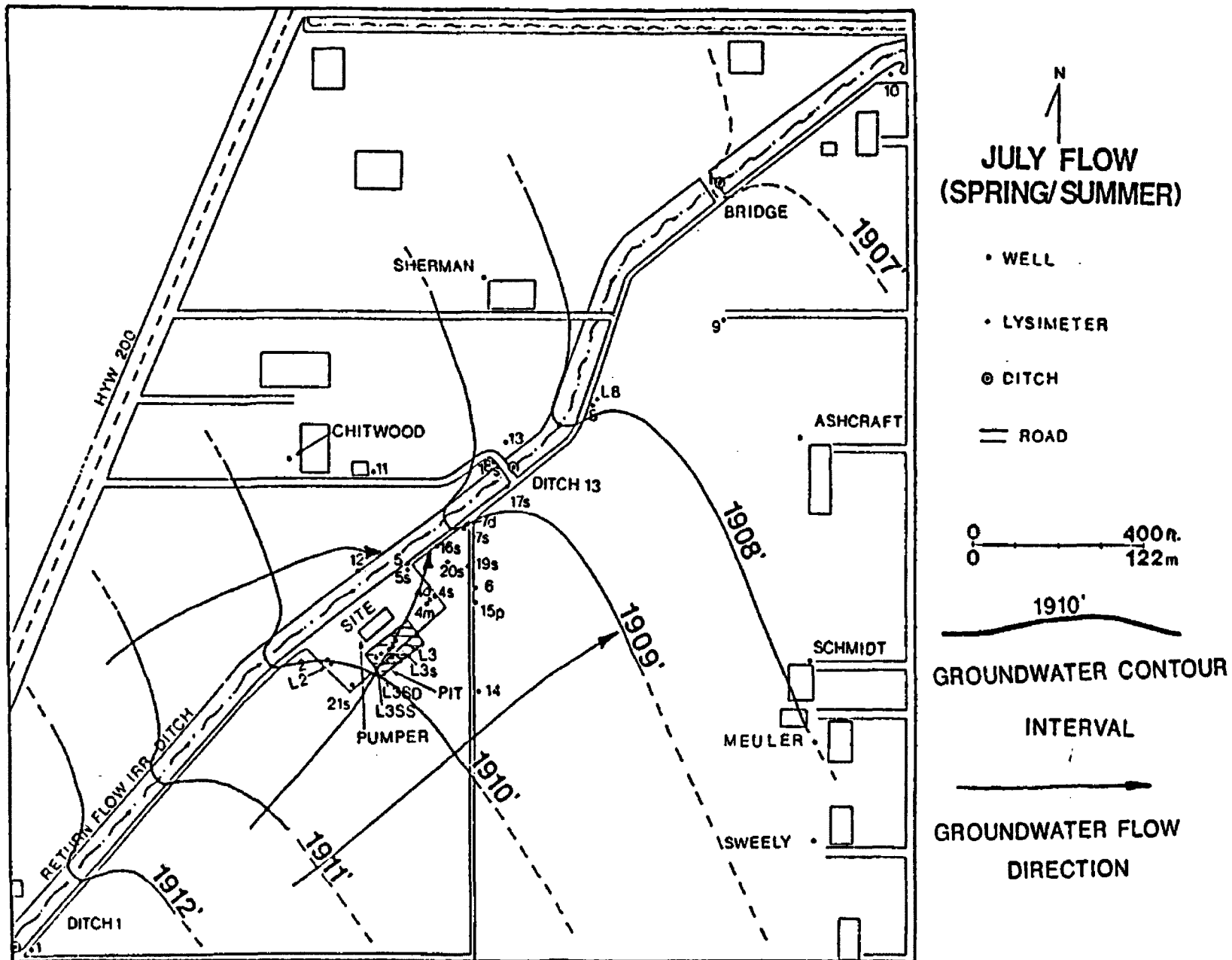


Figure 6. Groundwater flow map (spring and summer).

Figure 7 and Appendix G. Precipitation is lower in the spring and summer of 1988 compared to the spring and summer of 1987. In addition, evaporation is greater in 1988 compared to 1987. The higher evaporation and lower precipitation in 1988 indicate there was less surface recharge available to move pit contaminants compared to 1987. Since the reserve pit area is not flood irrigated, precipitation is the only surface recharge factor for leaching brine to the water table.

Annual water level fluctuation at well MW 3 is about four feet in the 1987-1988 year (Figure 8). The water table rise and fall is coincident with the start of spring flood irrigation in May and irrigation shutdown in November surrounding the site. The crops grown on the surrounding farmland (alfalfa, corn and sugar beets) require 18 to 24 inches (46 to 61 cm) of water during the growing season (SCS, 1987), which is significantly greater than the average annual precipitation of 14 inches/year (36 cm). The irrigation water used to supplement the crops is an important source of recharge to the shallow aquifer.

The peak 1988 water table position (Figure 8) matches the peak 1987 water table position, yet precipitation is less and free water evaporation is greater in 1988 compared to 1987. This suggests that more flood irrigation water was applied in 1988 compared to 1987 to counter drought conditions.

All monitoring wells and ditch locations have a similar annual water level fluctuation (Appendix H), indicating water

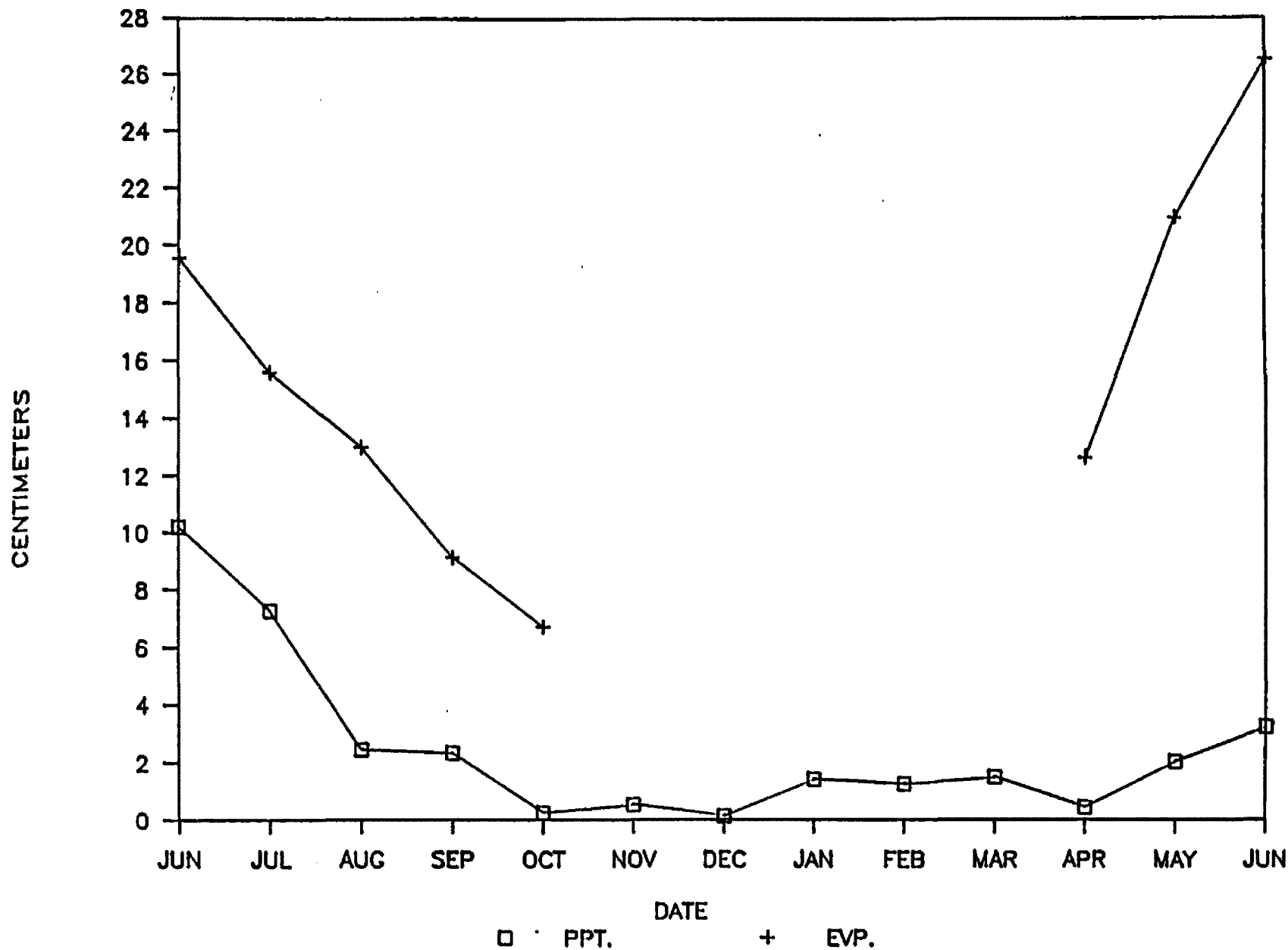


Figure 7. Annual precipitation and evaporation from June 1987 to June 1988 (Montana State University Agricultural Experiment Station Sidney, Montana).

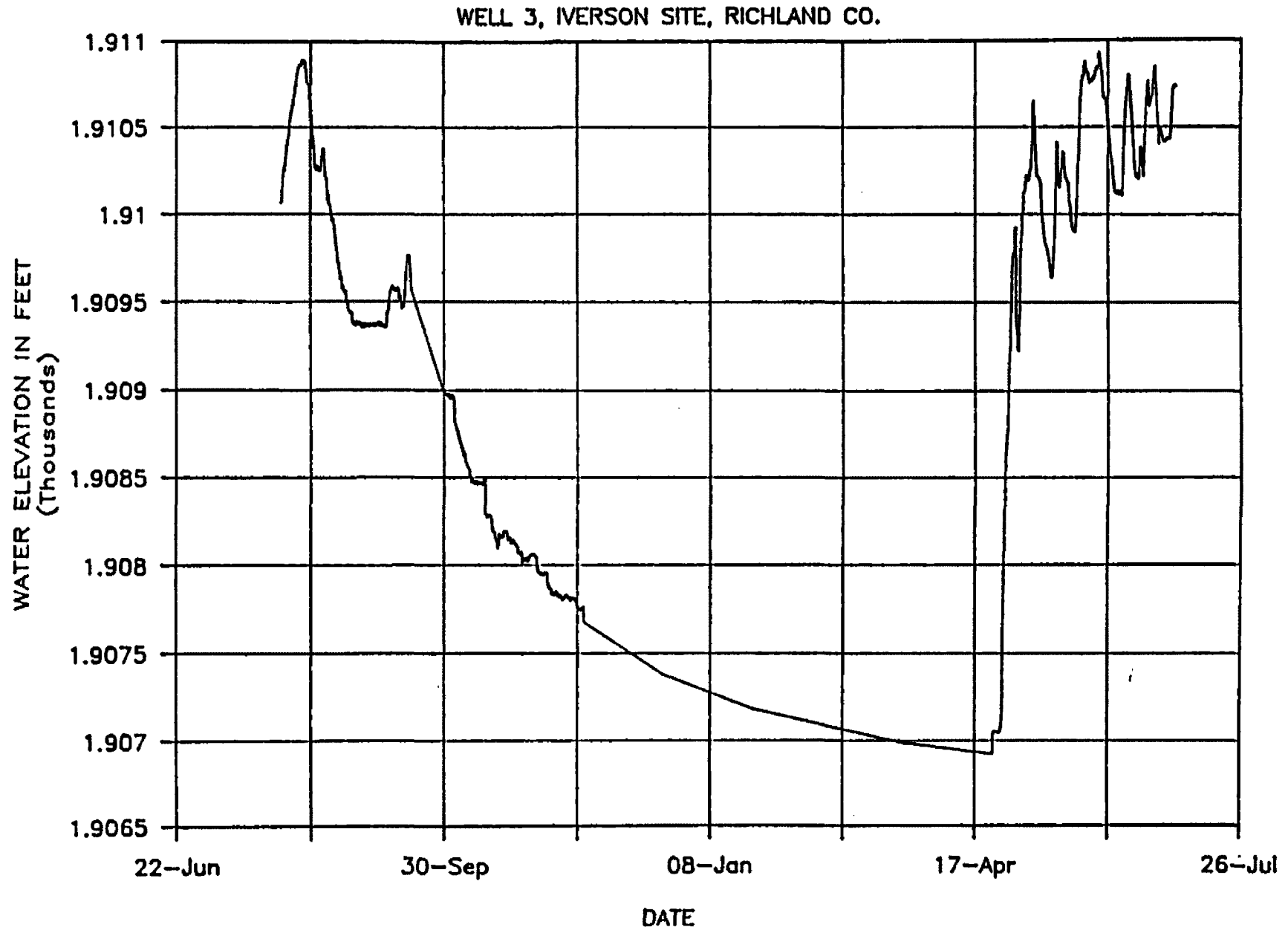


Figure 8. Annual water table fluctuation from August 1987 to July 1988.

level response to recharge is similar across the entire site.

Hydraulic conductivity and specific yield

One constant discharge aquifer test was used to estimate K (hydraulic conductivity) using both Boulton's partial penetration and delayed yield equations (Lohman, 1978). Delayed yield was evident in the pumping well (MW 15p) and one monitoring well (MW 6). The delayed yield method was employed to estimate K at these locations. Delayed yield was not evident in the other monitoring wells (MW 19s and MW 20s) and the partial penetration method was used to estimate K at these locations.

One foot of drawdown was measured at the end of the 4.5 hour aquifer test at the pumping well MW 15p (pumping 30 gpm), and only a few tenths of a foot drawdown were measured at the monitoring wells (Appendix I). Limited drawdown made K interpretation difficult. Hydraulic conductivity estimates ranged from 1,500 gal/day/ft² (7×10^{-2} cm/sec) to 25,000 gal/day/ft² (1.0 cm/sec) depending on the well location and whether the delayed yield or the partial penetration method was employed. K is probably close to 4200 gal/day/ft² (.2 cm/sec) since the most accurate simulated drawdown in an aquifer test was modeled using this value (to be discussed later). Assuming a 17.5 feet thick aquifer, transmissivity is 73,500 gallons/day/ft.

Hydraulic conductivity was also estimated by using the Theim steady-state equations (Driscoll, 1986) since drawdown

at the end of the aquifer test was small. Theim K estimates averaged one order of magnitude below K estimates calculated with the delayed yield and partial penetration methods. Transient K calculations probably represent a better estimate of K since steady-state conditions were not present at the end of the stress period.

A representative Sy (specific yield) for an unconfined sand and gravel aquifer was not calculated with the above methods. All calculated estimates were a number of orders of magnitude too low. Specific yield was estimated from 13 documented Sy values for sand and gravel aquifers (Turcan, 1963). Specific yield values ranged from 0.2 to 0.35, and .2 was chosen since it is a conservative estimate. Porosity of a repacked aquifer sample was determined to be .25 in the laboratory. To calculate groundwater velocity at the pit area (wells sites MW 3 and MW 7s) I used the following relationship:

$$V = \frac{K I}{7.48 n} \quad \text{where}$$

V = velocity (ft/day)
 K = hydraulic conductivity (4200 gal/day/ft²)
 I = gradient (.0026 to .001 depending on which month)
 n = porosity (.25)
 7.48 = conversion factor for gallons to cubic feet

and calculated the following monthly groundwater velocities:

87'	July:	5.6 ft/day	(1.70 m/day)
	August:	5.3 ft/day	(1.62 m/day)
	September:	4.7 ft/day	(1.43 m/day)
	October:	4.5 ft/day	(1.37 m/day)
	November:	3.2 ft/day	(.97 m/day)
	December:	2.9 ft/day	(.88 m/day)

88'	January:	2.9 ft/day	(.88 m/day)
	February:	2.8 ft/day	(.85 m/day)
	March:	2.6 ft/day	(.79 m/day)
	April:	2.4 ft/day	(.73 m/day)
	May:	5.6 ft/day	(1.70 m/day)
	June:	5.9 ft/day	(1.79 m/day)

Soil chemistry

Vadose zone soil analyses results are in Table 2 and Appendix K. Table 2 data represents the total acid digestible metal content for sodium, calcium and lead concentrations in samples collected from three reserve pit locations and three samples collected downgradient of the pit. Only sodium, calcium and lead are presented in the Table 2 since the other ions tested are at background concentrations. Chloride was not analyzed.

Table 2. Soil analyses (* less than detection)

<u>Sample site</u>	<u>Depth</u>	<u>% Na</u>	<u>% Ca</u>	<u>ppm Pb</u>
PITMUD 1	6-8 ft.	1.9	6.4	30.0
PITMUD 2	6-8 ft.	2.1	7.7	44.0
MW 3	5-7 ft.	1.6	3.8	21.9
MW 4	6-12 ft.	2.3	4.6	16.8
MW 7	2-7 ft.	0.8	2.6	*
MW 7 (DUPLICATE)	2-7 ft.	0.9	2.9	*
FT. UNION	28 ft.	0.7	7.8	*

Sodium, calcium, and lead results are elevated in pit samples PITMUD 1, PITMUD 2, and MW 3, and downgradient of the site at MW 4. Analyses indicate the pit mud results are approximately 50 to 75% greater than background analyses results (MW 7 and FT.UNION), with the exception of calcium in the Fort Union sample. MW 4 sample site is approximately 100 feet north of the pit area. The apparent soil contamination

at monitoring site MW 4 is probably from pit mud being pushed into a reclamation trenches extending from the reserve pit.

Groundwater and vadose water chemistry

Field measurements

Specific conductance (SC) and Quantab data (chloride concentrations) are in Figures 9 and 10, and Appendix J. Figure 9 shows Quantab and SC results for nested well site MW 4. The highest chloride and SC levels are present in upper four feet of the aquifer (MW 4s). Intermediate chloride and SC levels are present from four to eight feet below the water table (MW 4m). Background or near background chloride and SC levels are present in the lower half of the aquifer from eight feet below the water table to the top of the Fort Union Formation (MW 4d). Figure 9 also shows MW 4s chloride and SC levels in the summer of 1988 are less compared to chloride and SC levels in the summer of 1987.

Figure 10 shows Quantab and SC levels for monitoring wells MW 7s and MW 4m. Both wells are completed in the upper half of the shallow aquifer where brine contamination concentrates. Chloride and SC levels are significantly less in MW 7s than in MW 4m, and the peak timing of chloride and SC levels are approximately 2.5 to three months behind in MW 7s (which is down gradient of the MW 4 site). Contaminant loading in the spring and summer of 1988 appears to be less compared to the summer of 1987. Lower chloride and SC levels in 1988 are probably a result of less surface recharge and in

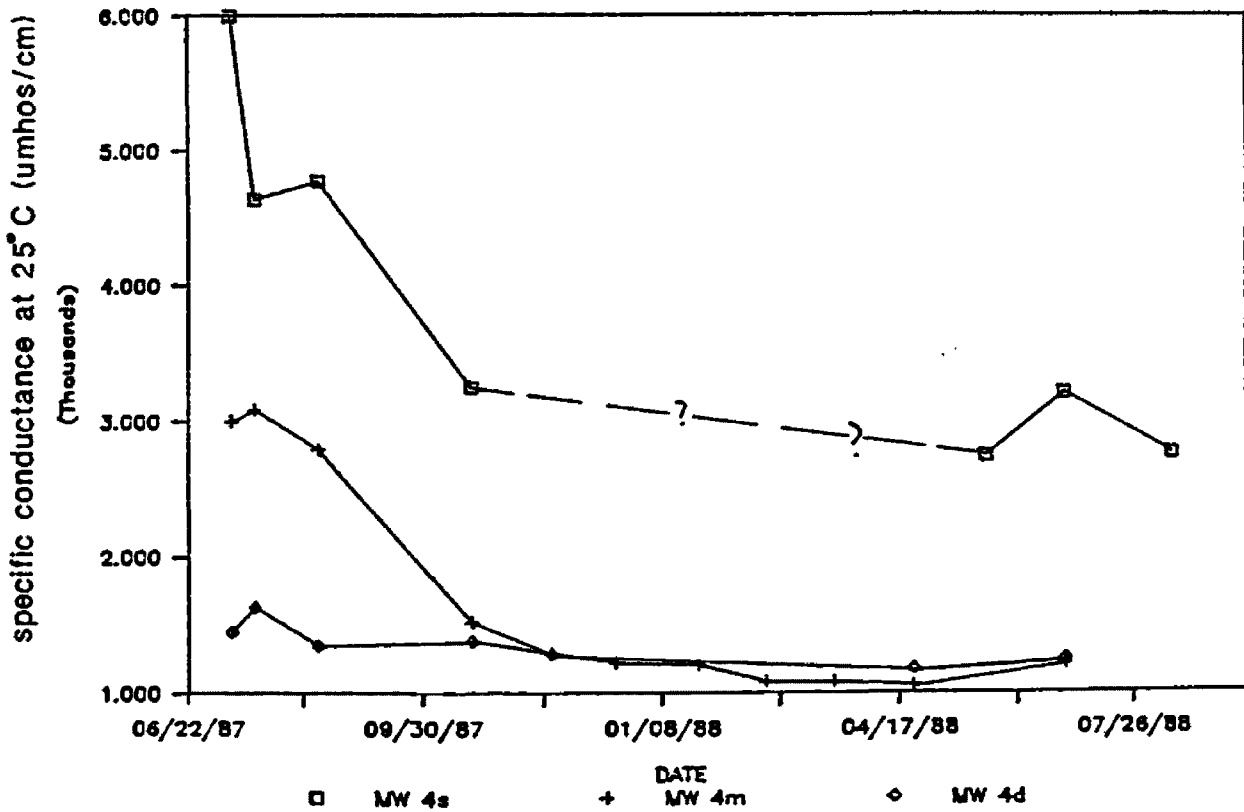
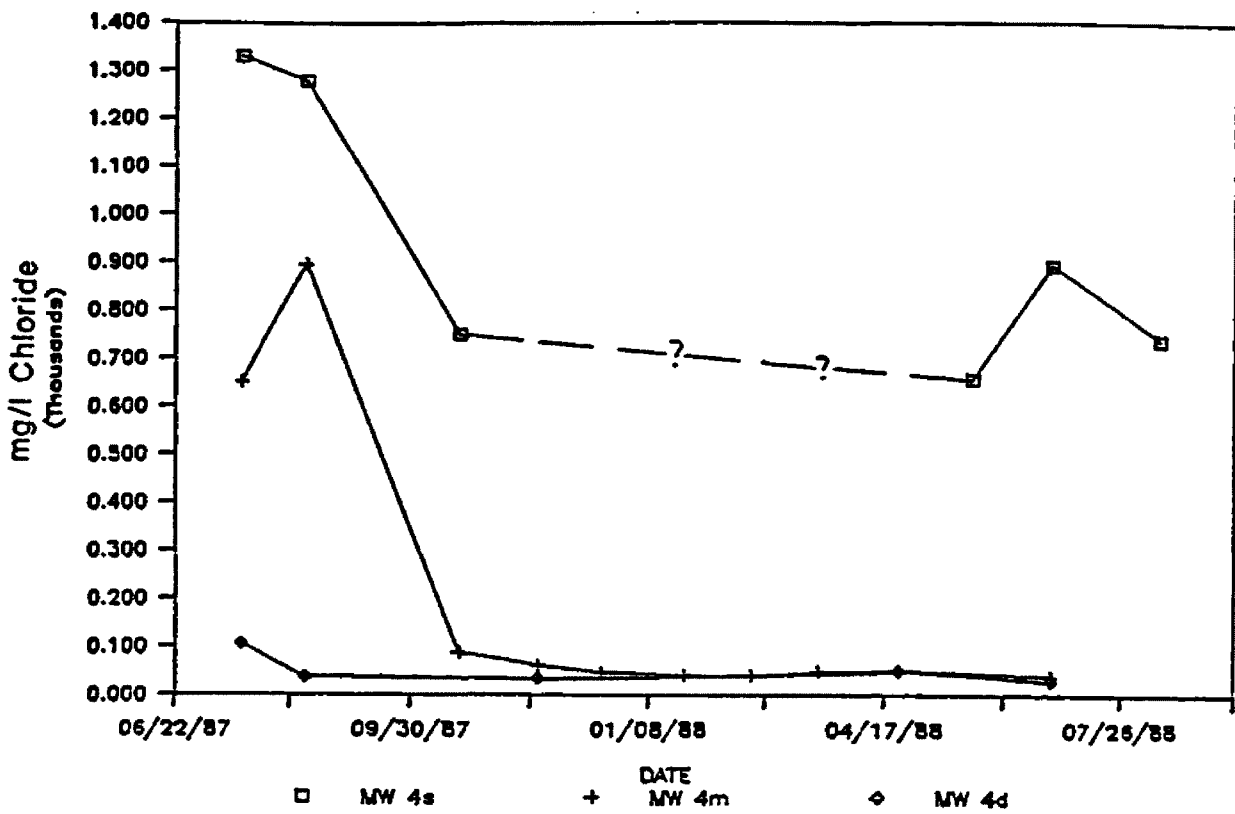


Figure 9. Annual chloride concentration and specific conductance at MW 4 nest.

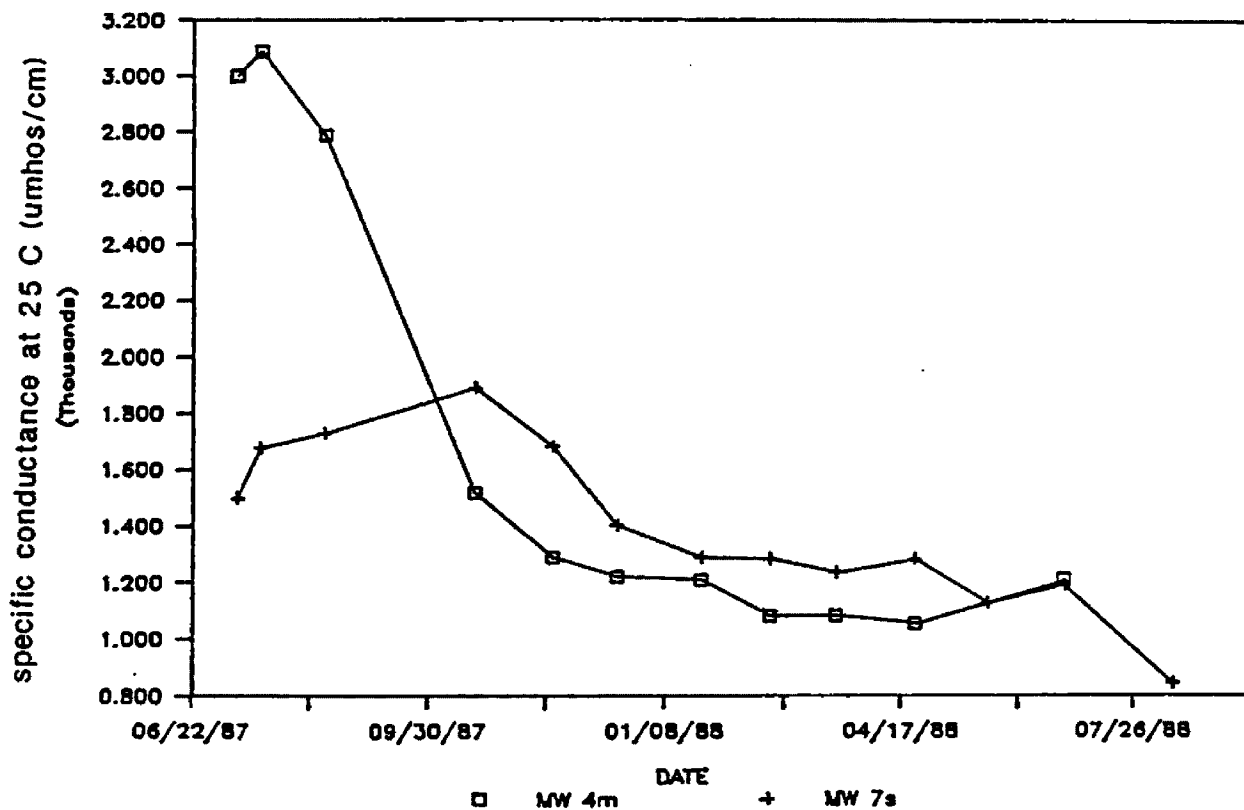
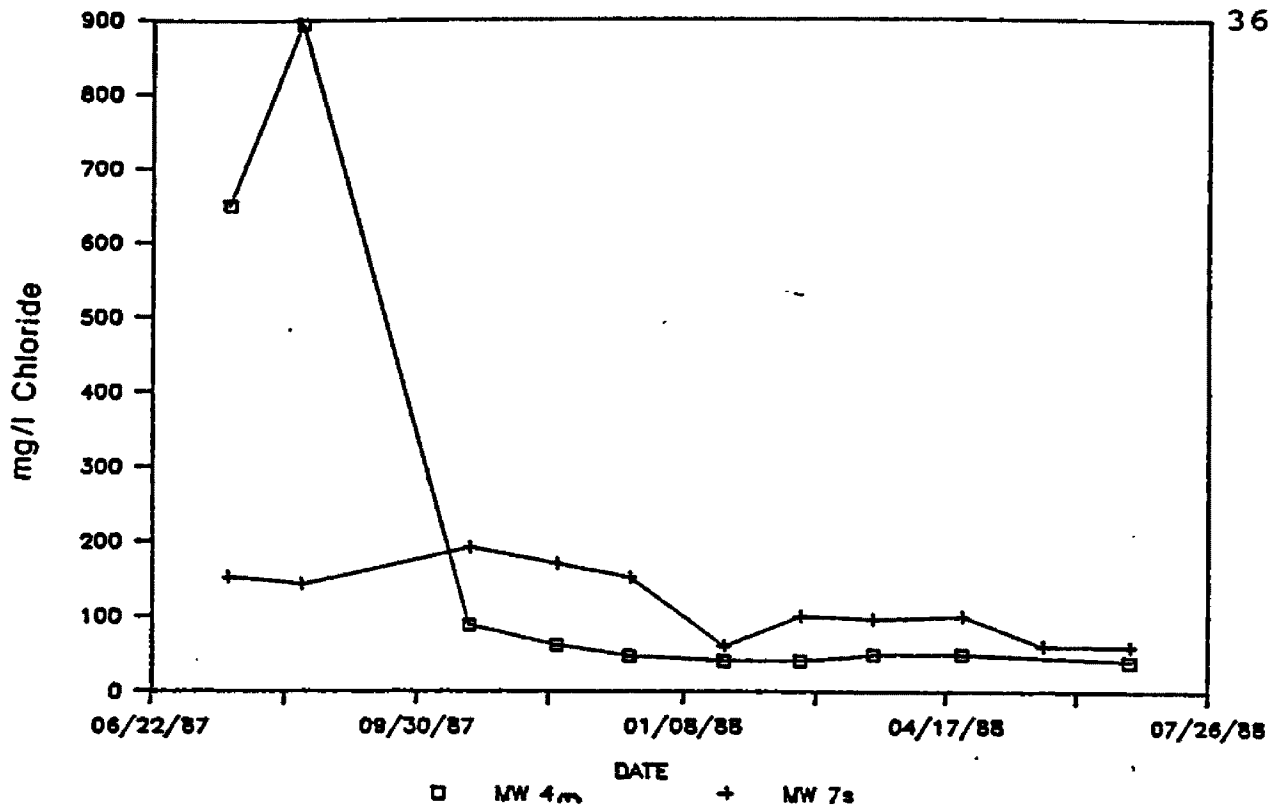


Figure 10. Annual chloride concentration and specific conductance at MW 4m and MW 7s.

turn less brine seepage into the underlying aquifer.

Field data indicate the brine contaminants tend to disperse near the top of the shallow aquifer. The lack of vertical contaminant migration suggests that K may be greater in the horizontal direction compared to the vertical direction and/or an upward flow component is keeping the pit contaminants near the top of the aquifer.

An approximation of groundwater velocity between the two well sites in Figure 10 indicates that contaminants move through the aquifer at approximately 3.3 ft/day to 4 ft/day. This was estimated by evaluating the time required for a chloride or SC peak at MW 4s to be measured at MW 7s. Figure 10 show a 2.5 to three month lag time between chloride and SC peaks which approximates the above velocities since the two sites are about 250 feet (80 m) apart.

Contaminants apparently enter the shallow aquifer by seepage from surface recharge and from contact between the water table and the pit mud. Monitoring well logs (Appendix D) show that pit mud extends to a depth of about nine feet (3 m) below land surface at the pit area. The water table fluctuates from seven to ten feet (2 to 3 m) below land surface at MW 3. This demonstrates that the groundwater intersects the pit material. The water table apparently saturates the pit mud and in turn pit contaminants are released into the aquifer.

In 1987, pit wastes probably entered the shallow aquifer

as a result of both seepage caused from precipitation and groundwater saturating the pit mud. Wetting fronts were less frequent in 1988 due to drought conditions, which apparently caused the lower chloride concentrations in the groundwater.

Laboratory analyses

Tables 3 and 4 show suspect brine contaminants in the shallow aquifer. Groundwater contamination is indicated by annual changes in ion and metal concentrations above background concentrations. This change is characterized by high spring and summer concentrations, lower fall concentrations and very low winter or background concentrations. Background concentrations are present under the pit area in the winter months indicating contaminant loading temporarily stops. Up gradient control wells (MW 2 and MW 21) have lower concentrations compared to wells in and downgradient of the reserve pit year round.

Pit soil moisture analyses at lysimeter L3 indicate the pit material is contaminated with all metals and ions tested (except titanium and zinc) at concentrations sometimes a number of orders of magnitude above those in the groundwater.

Chloride and sodium are the two most elevated constituents detected in the reserve pit mud and groundwater. Other ions and metals were detected in the pit mud and groundwater but to a lesser degree compared to chloride and sodium. Constituents detected above the recommended federal drinking water standards are nitrates (>10 mg/l), Manganese

Table 3. Ions in the shallow aquifer attributed to brine seepage (dissolved ions reported as mg/l)*.

PARAMETER	DATE	LOCATION				
		SITE 2/21	SITE 3	SITE 4	MW 7s	L3
Ca	07/22/87	145	357	226	119	9510
	10/22/87	111	113	172	109	
	04/23/88	102	96.3	83.2	86.4	8350
	06/28/88	105	147	175	84.9	7310
Mg	07/22/87	31.2	90.7	50.6	53.6	1270
	10/22/87	34	35.8	44.5	53.2	
	04/23/88	32.4	37.4	42.6	42	1260
	06/28/88	37.2	38.7	43	43.3	1300
Na	07/22/87	63.5	1436	681	196	8470
	10/22/87	85.2	229	462	211	
	04/23/88	91.8	105	143	162	88600
	06/28/88	89.1	324	476	146	73500
K	07/22/87	3.8	25.4	28.4	8.2	1040
	10/22/87	4.5	8.1	18.7	8.8	
	04/23/88	3.3	4.5	0.6	6.3	1020
	06/28/88	5	14.4	19.2	7.3	885
Cl	07/22/87	31.9	2660	1060	217	14500
	10/22/87	19.7	221	678	224	
	04/23/88	21.1	24.4	28.2	70.7	156000
	06/28/88	23.2	411	680	60.6	130000
Mn	07/22/87	<.001	5.01	0.29	0.003	42
	10/22/87	<.001	0.4	0.085	<.001	
	04/23/88	0.006	0.28	0.022	0.015	35.9
	06/28/88	0.001	0.077	0.001	0.002	30.9
HC03	07/22/87	349	532	589	476	220.8
	10/22/87	438	484	539	478	
	04/23/88	438	470	501	506	332
	06/28/88	382	441	508	448	268.4
SO4	07/22/87	203	256	253	243	690
	10/22/87	187	208	227	227	
	04/23/88	213	233	251	267	1110
	06/28/88	209	243	260	234	223
N	07/22/87	22.7	2.61	9.14	4.03	25
	10/22/87	5.89	3.85	3.21	2.22	
	04/23/88	4.34	1.55	0.97	1.24	<.04
	06/28/88	11	10.6	3.28	2.75	36

*See also Br Appendix L

Table 4. Metals in the shallow aquifer attributed to brine seepage (dissolved metals reported as ug/l).

PARAMETER	DATE	LOCATION						L3
		MW 21s	MW 3	MW 3s	MW 4s	MW 16s	MW 20s	
B	07/21/87		1030					
	10/22/87		220				570	
	04/23/88		140					38000
	06/28/88	120		400	820	600		38000
Li	07/21/87		370					
	10/22/87		71				160	
	04/23/88		20					13100
	06/28/88	20		130	190	100		10000
Ba	07/21/87		530					
	10/22/87						92	
	04/23/88		60					320
	06/28/88			110	100			270
Sr	07/21/87		7300					
	10/22/87		1020				2510	
	04/23/88		990					241000
	06/28/88	630		2660	2780	940		207000
Ti	07/21/87		29					
	10/22/87		3				5	
	04/23/88		<1					<1
	06/28/88	4		6	8	5		<1
Zn	07/21/87		14					
	10/22/87		4				<3	
	04/23/88		<3					880
	06/28/88	<3		<4	3	<3		430
Be	07/21/87		21					
	10/22/87							
	04/23/88		<1					26
	06/28/88			1				21

(>0.05 mg/l) and chloride (>250 mg/l). Nitrates and Manganese only exceed these standards in the spring and summer at the pit area; chloride is generally above the recommended drinking water standards year round.

Ions and metals at elevated concentrations in the pit mud but not detected in the groundwater are probably absorbed, bonded, precipitated and/or adsorbed by the clay minerals of the pit material. An x-ray diffraction pattern of the pit mud (Appendix M) shows the drilling mud is rich in smectite (or bentonite, which is the common drilling mud used drill oil wells). Smectite is a 2:1 layer silicate clay mineral that will absorb hydrated ions in the interlayer spaces of the 2:1 structure (depending on the size and charge of the element or compound).

The ions that do not seep into the groundwater are identified by comparing the pit mud soil moisture analyses with the groundwater analyses. They are: nickel, vanadium, zirconium, aluminum, silver, cadmium, iron, chromium, copper, molybdenum, and lead (Appendix K and L). These are retained in the pit mud.

Figures 11 through 14 are chloride plume maps. Most contaminant loading occurs in the spring and summer as shown by high chloride concentrations originating from the pit area in Figures 11 and 12. Loading wanes in the fall and chloride concentrations drop (Figure 13). In the winter, the water table continues to lower and little surface recharge is

CI CONCENTRATION mg/l FOR JULY, 1987

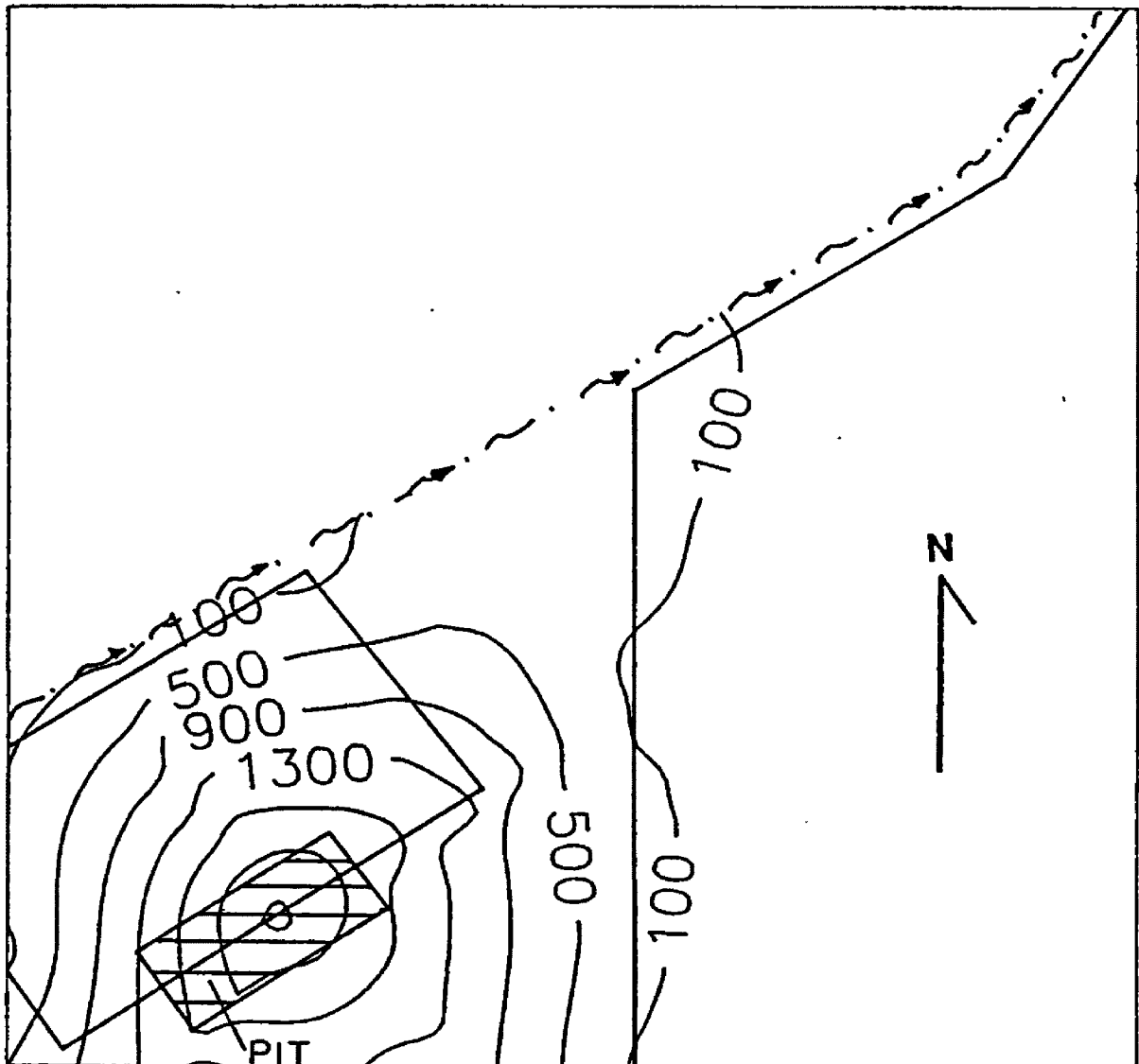


Figure 11. Chloride plume map July 1987 (CI = 400 mg/l).

Cl CONCENTRATION mg/l FOR OCT., 1987

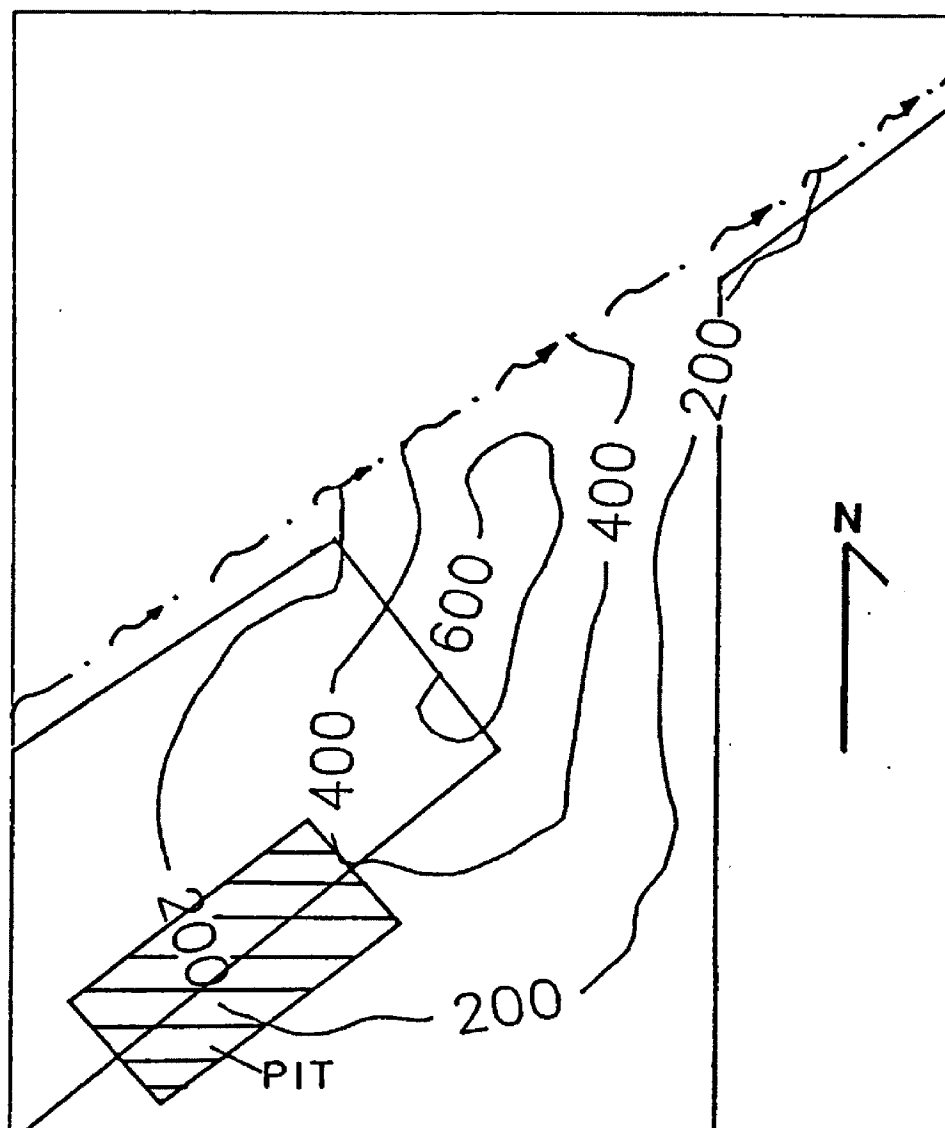


Figure 12. Chloride plume map October 1987 (Cl = 200 mg/l).

Cl CONCENTRATION mg/l: APRIL, 1988

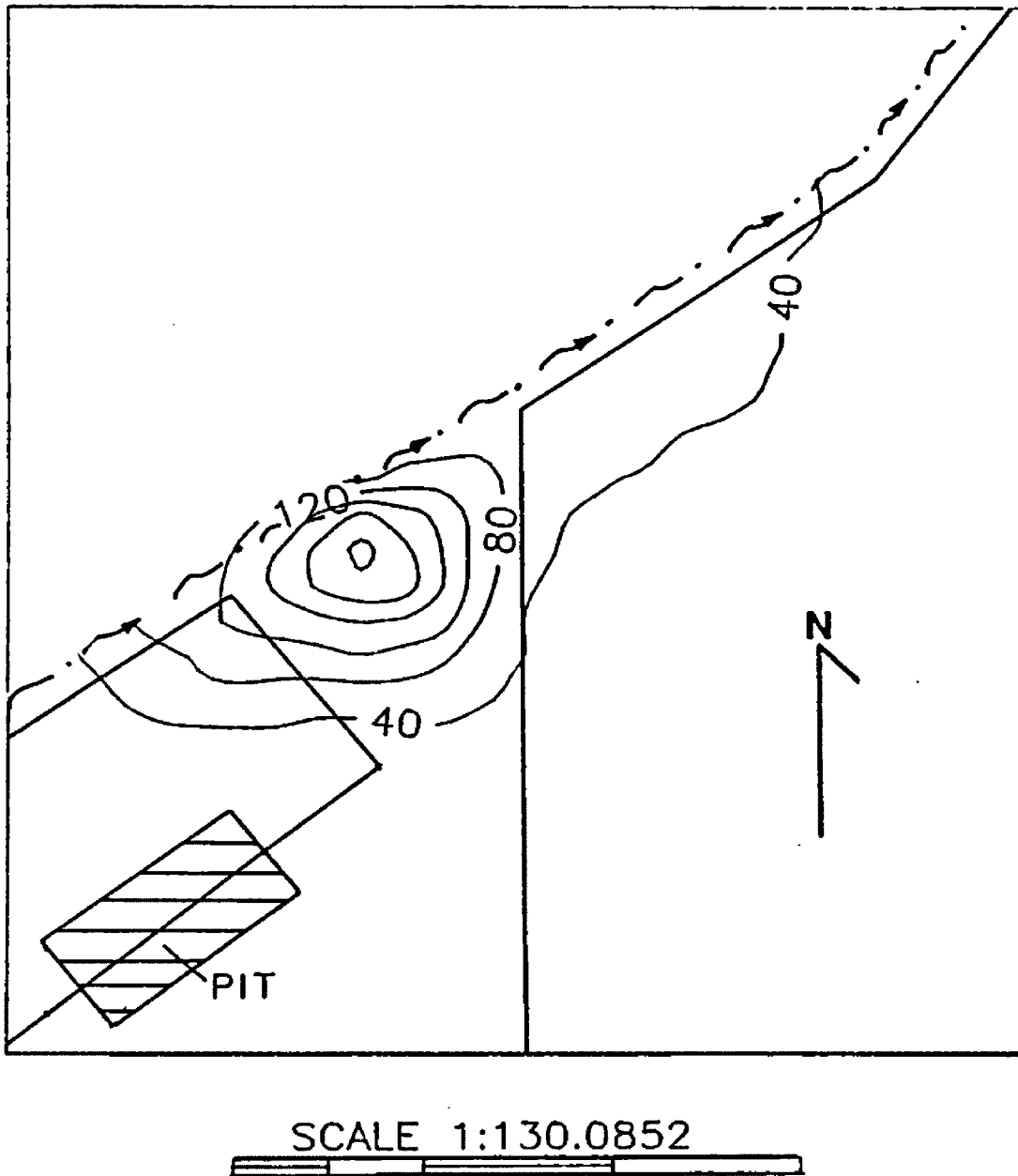


Figure 13. Chloride plume map April 1988 (Cl = 40 mg/l).

Cl CONCENTRATION mg/l: JUNE, 1988

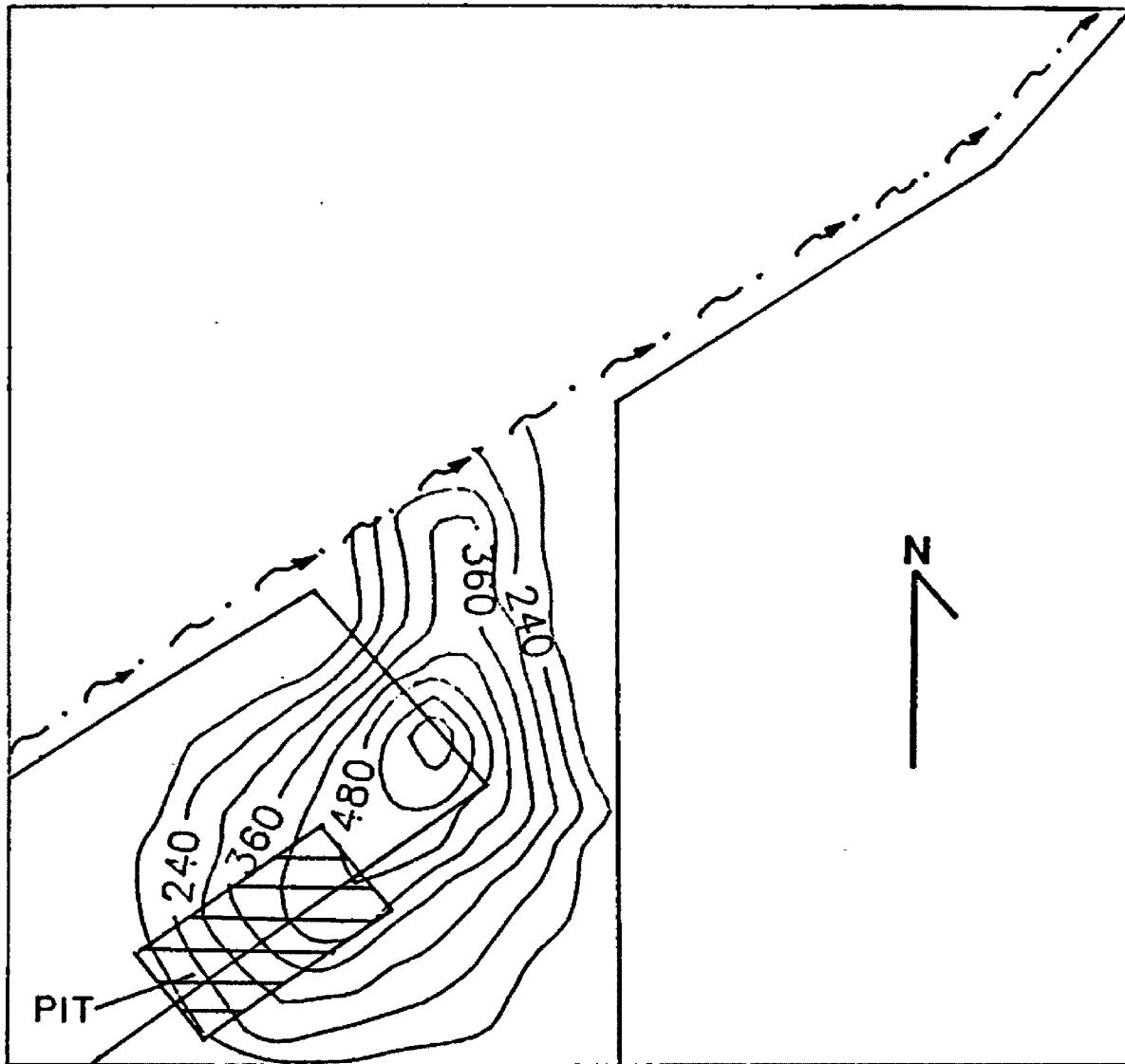


Figure 14. Chloride plume map June 1988 (Cl = 60 mg/l),

available. Background chloride concentrations are present in the groundwater under the reserve pit at this time (Figure 14). In the spring, chloride concentrations increase at the pit area from the high water table intersecting the bottom of the pit and increased precipitation leaching pit contaminants into the groundwater.

Shallow groundwater discharges to the irrigation ditch in the spring, summer and fall since the ditch acts as a groundwater sink, removing chloride from the aquifer. In late winter, the ditch is influent and a small groundwater divide probably forms under the ditch causing chloride migration to be parallel to the ditch.

Figures 11 through 14 and Tables 3 and 4 show that contaminants load into the shallow aquifer in an annual cyclic fashion. In addition, Figures 11 through 14 show that the chloride concentrations decline dramatically over a short distance from the source (350 feet or 115 m), from 2800 mg/l under the pit to 600 mg/l at MW 16s.

Data plots for Figures 11 through 14 are in Appendix N, which also contains SC plume maps and data for the same locations and seasons.

Michelle Dewey of the University of Montana studied the same reserve pit site from 1982 to 1983. Dewey completed five monitoring wells in the shallow aquifer and collected water quality samples for major ion analysis. Chloride concentrations down gradient of the reserve pit at Dewey's

wells 3 and 5 (Appendix O) have similar chloride concentrations as down gradient monitoring wells sampled in this study. Background chloride concentrations at Dewey's well 1 (Appendix O) also has similar chloride concentrations as background wells sampled in this study (with the exception of one anomalous peak in April, 1983).

A comparison of Dewey's data and data collected in this study indicate that brine seepage has not decreased significantly in the past four years, suggesting the pit will be a long term contaminant source. However, comparison of Dewey's data with this study is awkward since her wells were located at different points and completed differently. The comparison does show the pit mud on site may load brine wastes into the shallow aquifer for the next tens of years.

Yellowstone River chemistry

Data provide by the United States Geological Survey shows no significant increase in chloride, sodium, sulfate, calcium, potassium, and magnesium concentrations occur in the chemistry of the Yellowstone River between Terry, Montana and Sidney, Montana (Slagle, 1983, 1984 and 1988) (Appendix O). Terry is approximately 90 miles (155 km) up stream of Sidney on the Yellowstone River. Many reserve pits are located along this stretch of the Yellowstone River in both valley fluvial deposits and the nearby Tertiary deposits. Brine from these reserve pits may seep into the underlying aquifers and in turn increase chloride and other ion concentrations in the

Yellowstone River (assuming groundwater velocity, time, dispersion and discharge will move the contaminants to the river). Chemical data collected from the Yellowstone River indicate dilution of brine contaminants in the groundwater or in the Yellowstone River are too great to measure an increase in concentration of ions analyzed or no increase occurs.

Apparent Conductivity and resistivity

EM

Results of the EM 10 m horizontal and vertical mode surveys are in Figures 15 and 16. The EM 20 m and 40 m surveys show similar conductivities to those measured in the EM 10 m survey (Appendix P). All EM data were collected in July, 1987. Figures 15 and 16 show soil conductivity data measured near the reserve pit and do not include the conductance data measured over the entire EM 10 m grid in Figure 5. EM 10 m conductance data collected away from the pit were background readings and are plotted in Appendix P.

In Figure 15, the pit area is delineated by conductance values ranging from 40 to 110 mmhos/m compared to background readings of 20 to 25 mmhos/m. Only a limited area around the pit shows a high conductivity. Figure 16 accurately defines the pit area with conductivity values ranging from 90 to 160 mmhos/m, and appears to delineate a conductive zone (30 to 60 mmhos/m) in the shape of a plume downgradient of the pit, similar to the chloride plume map in Figure 14.

Water quality data for July, 1987 indicate that brine

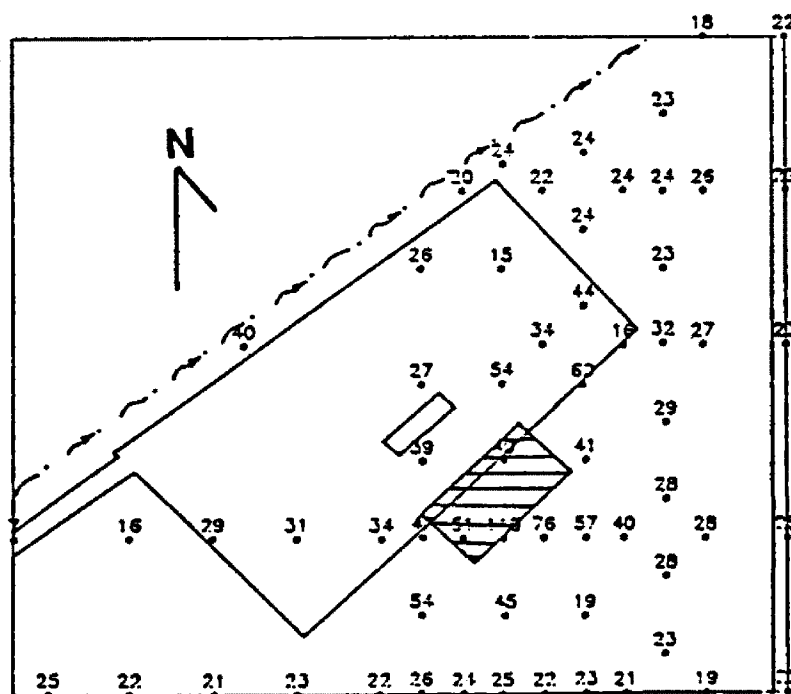
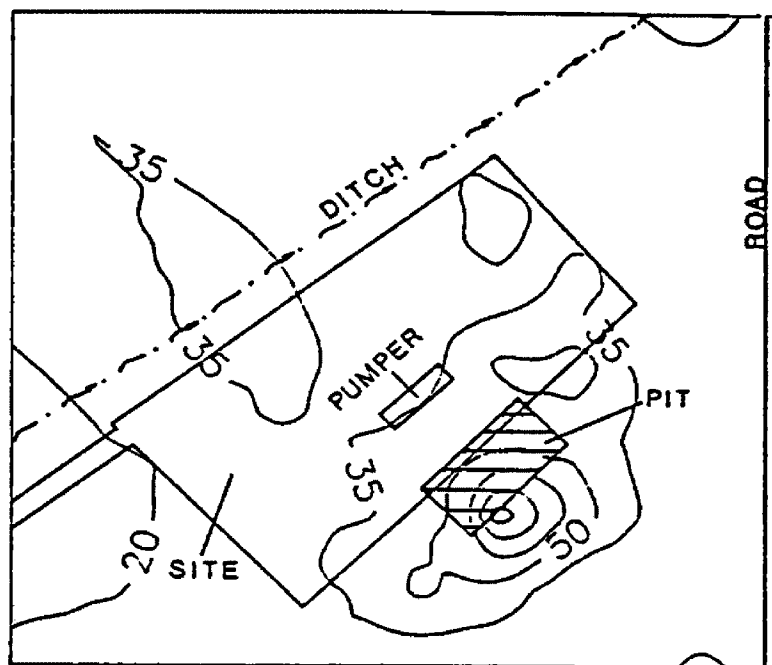


Figure 15. Contour map and data plot of EM 10m horizontal mode survey (CI = 15 mmhos/m) (July 1987).

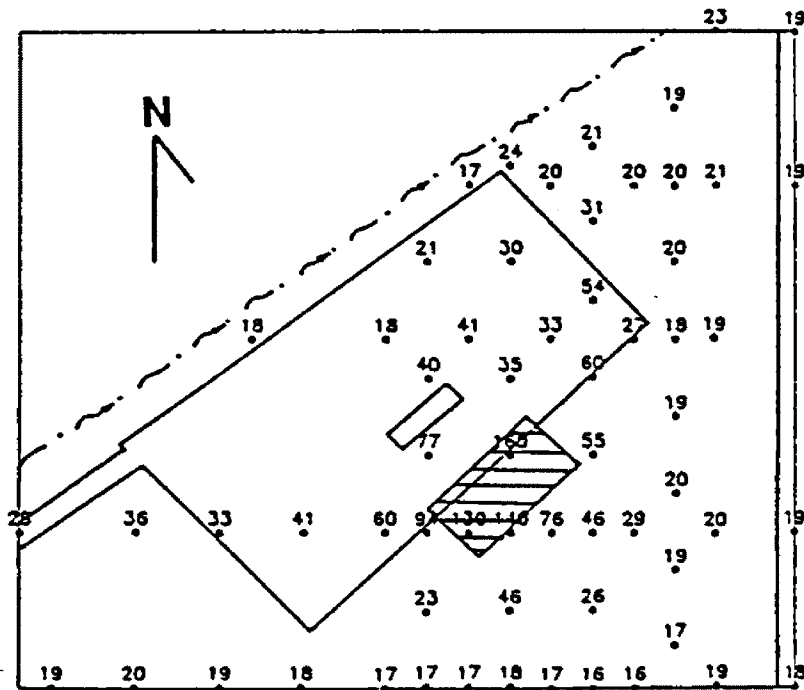
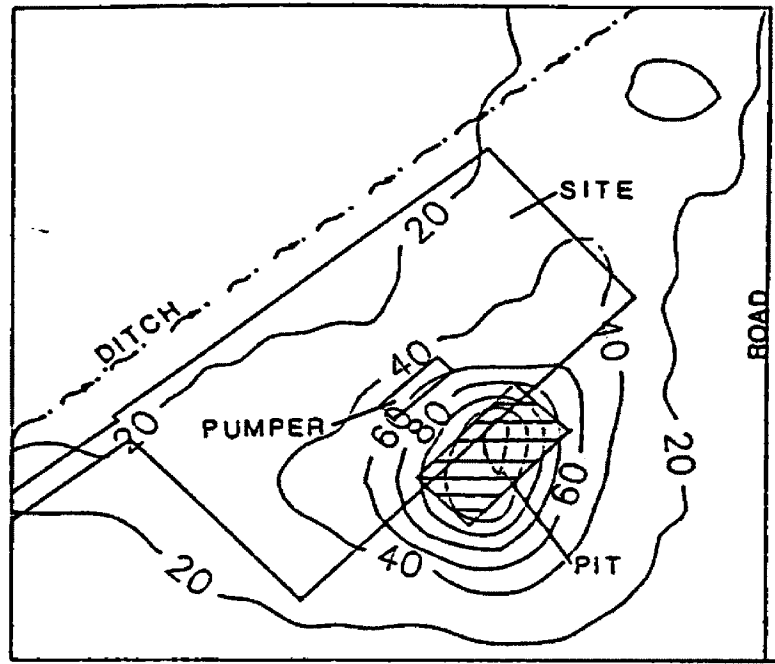


Figure 16. Contour map and data plot of EM 10m vertical mode survey (CI = 20 mmhos/m) (July 1987).

contamination was present downgradient of reserve pit. The presence of brine in the groundwater was apparently significant enough that the EM 10 m vertical mode could delineate the contaminated area. The EM survey may also have delineated soil contamination. This is indicated from lysimeter data at L2 which shows contaminated soil moisture approximately four to six feet below ground surface, 100 feet (30 m) west of the pit area. Soil chemistry data at the MW 4 site indicates that contaminated pit mud is present from approximately six to 12 feet (4 m) below land surface, 100 feet (30 m) north of the pit area.

The contrast between the EM data in Figures 15 and 16 is probably caused by a difference in EM depth penetration. The horizontal mode (Figure 15) is less influenced by the surface and near surface conductivity, probing soil approximately 1.5 times the coil spacing. The vertical EM mode (Figure 16) examines conductivity from the ground surface down to about .75 times the coil spacing. The shallower EM penetration (vertical mode) did a better job delineating brine contamination because the water table and reserve pit are within seven to nine feet of the ground surface. The "deeper" soil was not as affected by the brine contamination making the vertical EM mode a better tool to evaluate brine contamination at this site.

A comparison of EM conductivity values with chloride and SC levels collected in July 1987 show the EM highs are located

in the same area as chloride and SC highs (Appendix M). A correlation of EM versus chloride or SC could not be determined with these data. Monthly EM data, collected during water quality sampling, is needed to quantitatively determine groundwater quality from EM conductivity.

Results indicate that EM is a valuable tool for locating brine contamination. Highly conductive pit mud and brine water are delineated with the equipment.

Resistivity

Resistivity 10, 20 and 30 foot a-spacing data and plots support conductivity data collected with EM (Appendix N). Low resistivity values are present in the pit area extending northward in the same direction as groundwater flow away from the pit area. Operation of resistivity equipment was significantly more time consuming and cumbersome to use compared to the EM equipment.

Resistivity data collected in this study are very similar to Dewey's resistivity work in 1983, suggesting that contaminant concentrations have not significantly changed since the first resistivity surveys in 1983.

Groundwater and solute transport modeling

Flow modeling

Figure 17 displays the model area. Figure 18 displays the node location, boundary conditions, and site landmarks. K was set at 4200 gal/day/ft² for each node except those representing the irrigation ditch. The bottom of the ditch

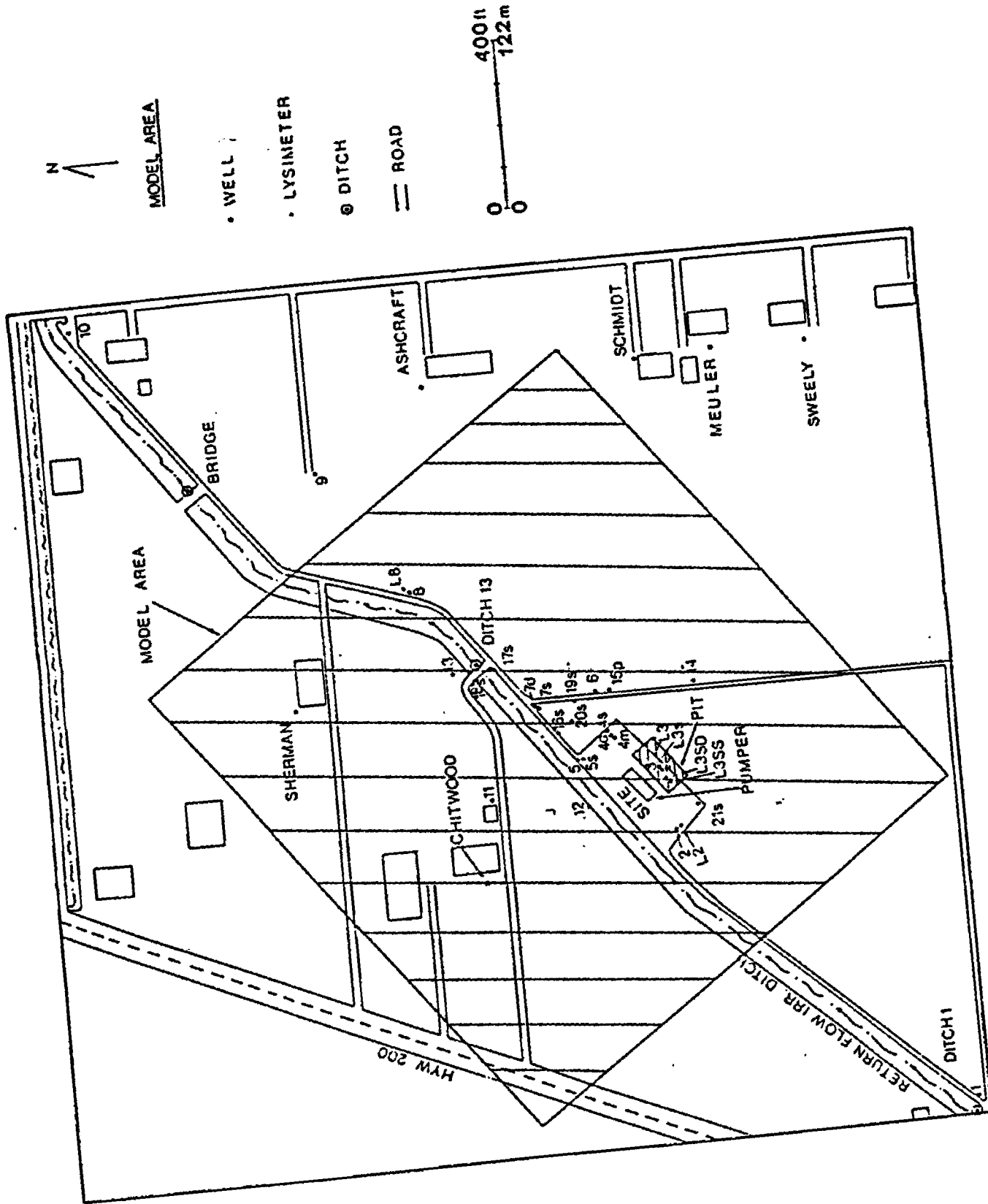
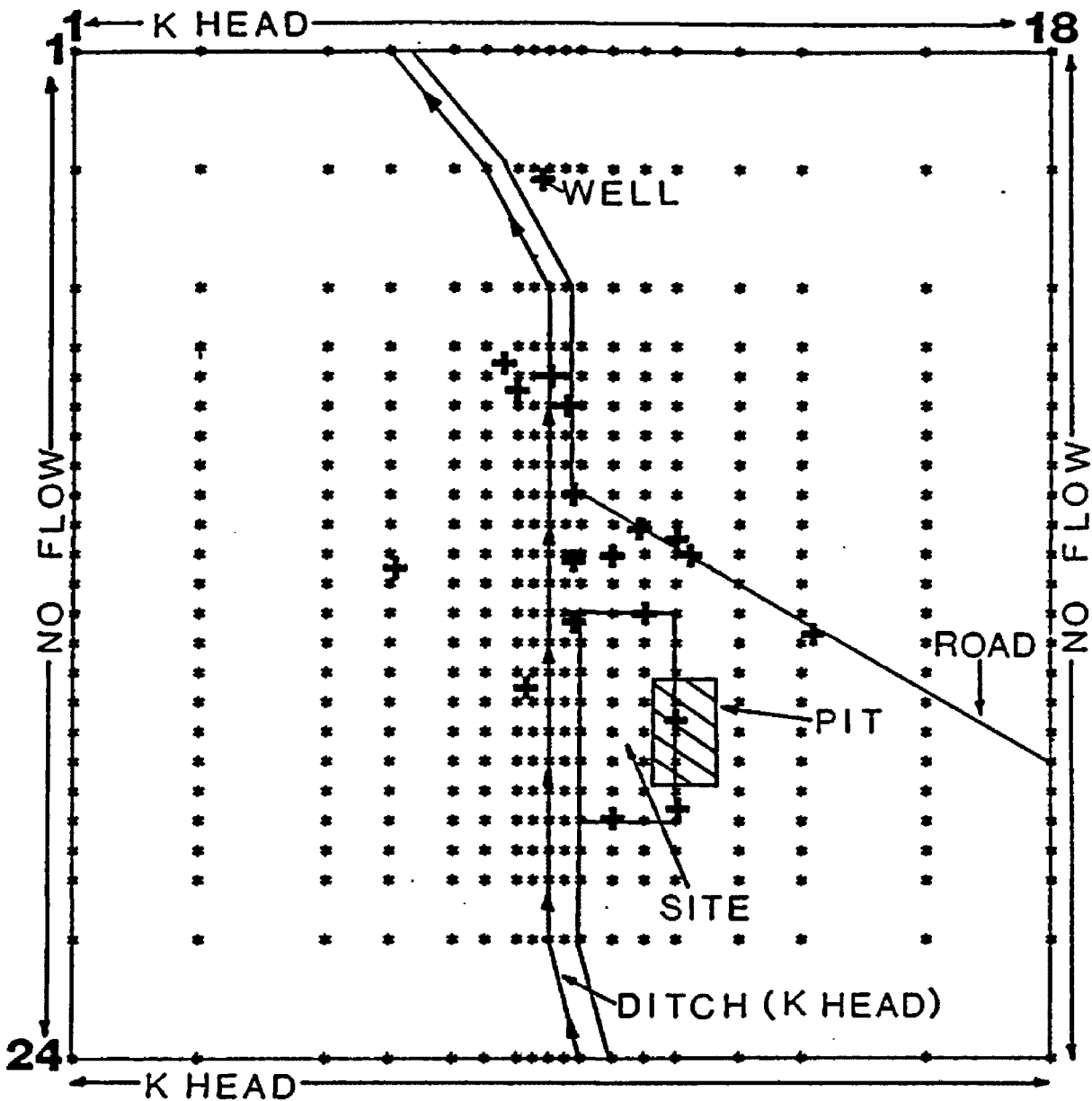


Figure 17. Model area.



SCALE : 0 100 ft
 0 30 m

KEY

- K-HEAD = constant head boundary
- + = well location
- NO FLOW = no flow boundary

Figure 18. Boundary conditions.

is composed of mud and silt, and consequently is assumed to be less transmissive than the underlying sand and gravel. The actual K of the ditch bottom is unknown but probably is a number of orders of magnitude lower than the sand and gravel aquifer. Since PLASM is a two-dimensional model it interprets K to penetrate the entire aquifer thickness. To account for the lower K in the ditch, K was lowered in all ditch nodes to 420 gal/day/ft² which approximates the actual ditch K by averaging K in the ditch bottom and the aquifer. No other K zones were modeled. Specific yield was set at .2 for all nodes.

October, 1987 head data were used to calibrate a steady-state flow model simulation. Ten time steps were used, with a model error tolerance of .1 feet. Constant head data were varied until simulated head were within .1 to .2 feet of actual head data (see October data Appendix R and Appendix S).

Modeling groundwater flow alone, a wide range of K inputs could be used to simulate head within the .1 to .2 error tolerance using the same constant head data. To determine a more precise estimate of K, the aquifer test was simulated. In order to match the actual and simulated aquifer test drawdown, K had to be limited to approximately 3500 to 4500 gallons/day/ft² and Sy equal to .2 (Appendix T).

Simulated drawdown at the pumping well is off by approximately one foot. This is caused by the model removing water from a nodal area instead of discrete well diameter,

which in turn caused the simulated drawdown at the pumping node to be less than actual well drawdown.

After achieving a steady-state simulation by adjusting constant head data and simulating the aquifer test, the same model and different constant head data were used to simulate 12 months of transient head data. Each month was modeled by reading a new constant head file (Appendix S) every thirty days and adjusting the constant head files until head was simulated within a .2 to .3 foot error tolerance.

An error tolerance of .2 to .3 feet was chosen after considering factors that might account for inaccuracies in the actual head data. These are: human error, survey elevations and accuracy of well location in the model.

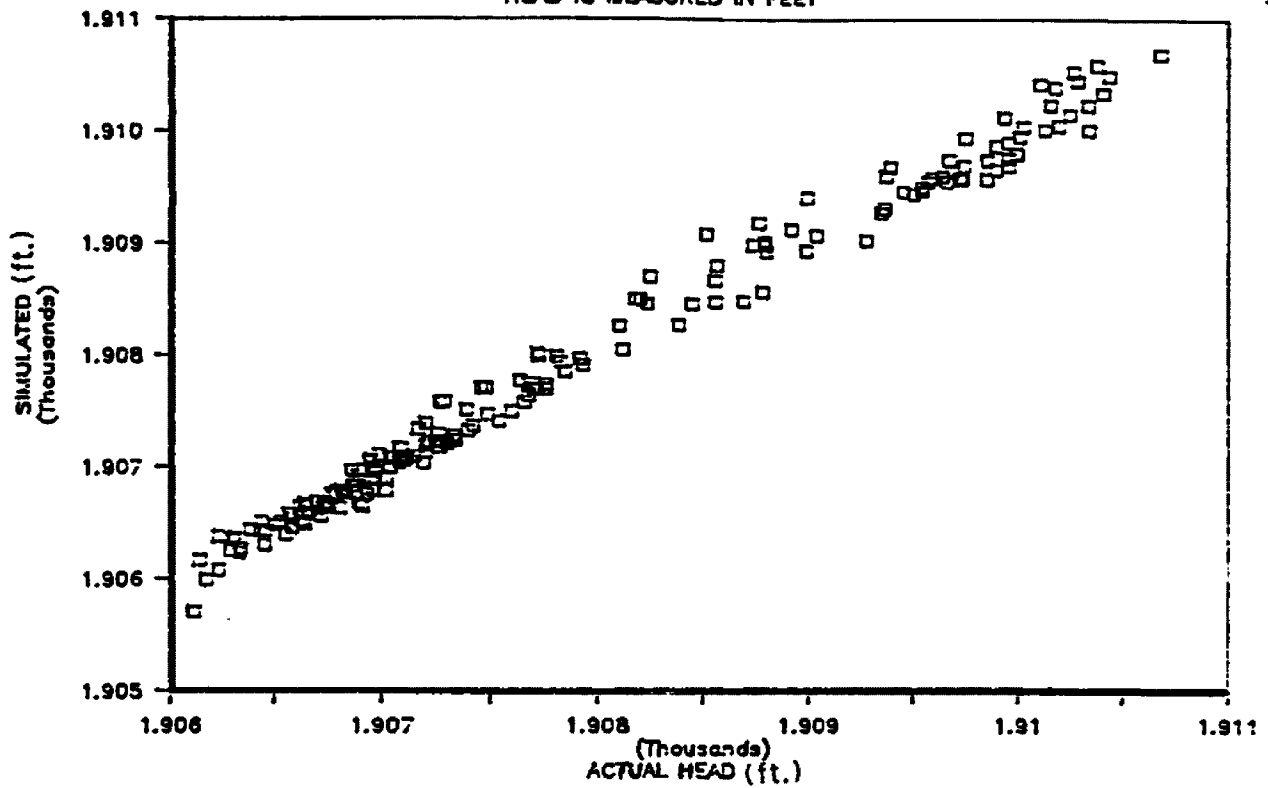
Simulated monthly groundwater flow from July 1987 to June 1988 are in Appendix R. The accuracy of each monthly flow field is depicted in Figures 19 through 21 and Appendix T. Figure 19 is a regression plot of the actual and simulated head from both the transient (top diagram) and steady-state simulations (bottom diagram). Both regressions show a good correlation between the actual and simulated head.

Figure 20 compares the actual and simulated head in monitoring wells MW 3 and MW 7. This figure shows a good correlation between actual and simulated head for one year of transient modeling.

Figure 21 is a variance plot of simulated head data. This figure shows that 85% of the simulated head are within

TRANSIENT MODEL REGRESSION

HEAD IS MEASURED IN FEET



STEADY-STATE MODEL REGRESSION

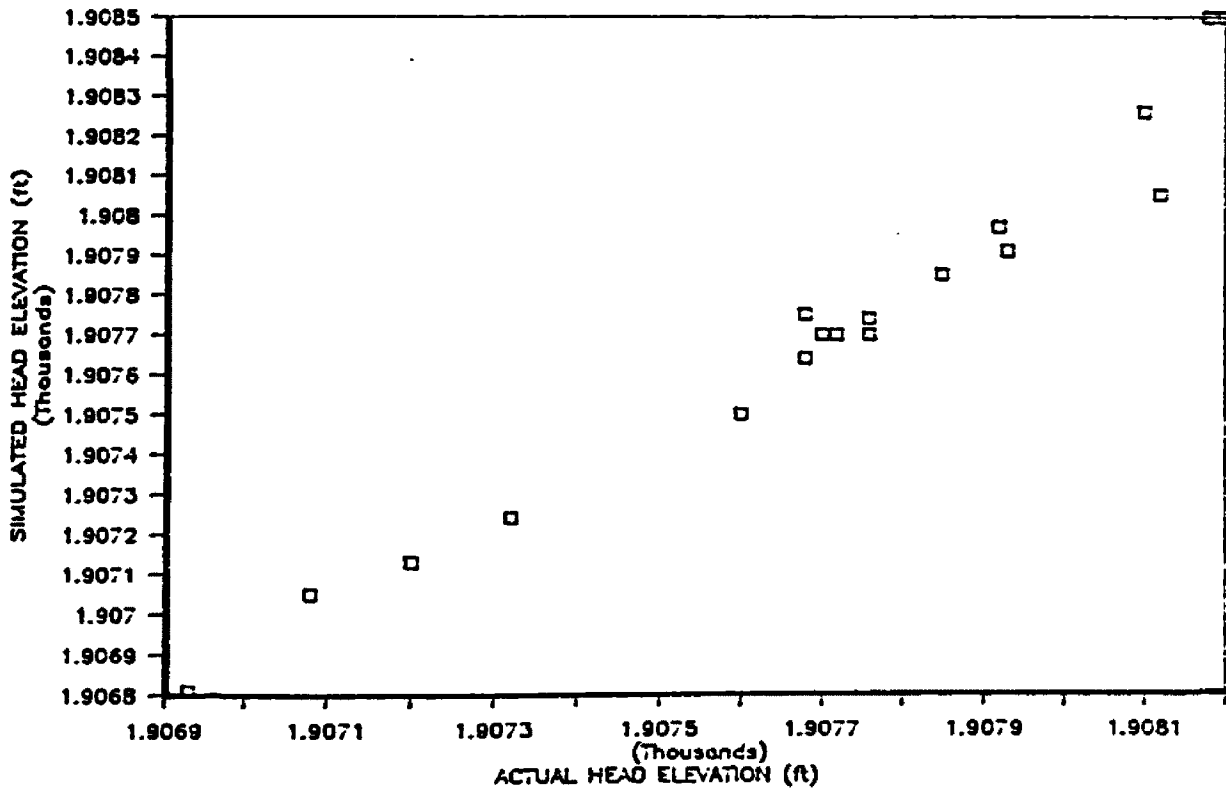
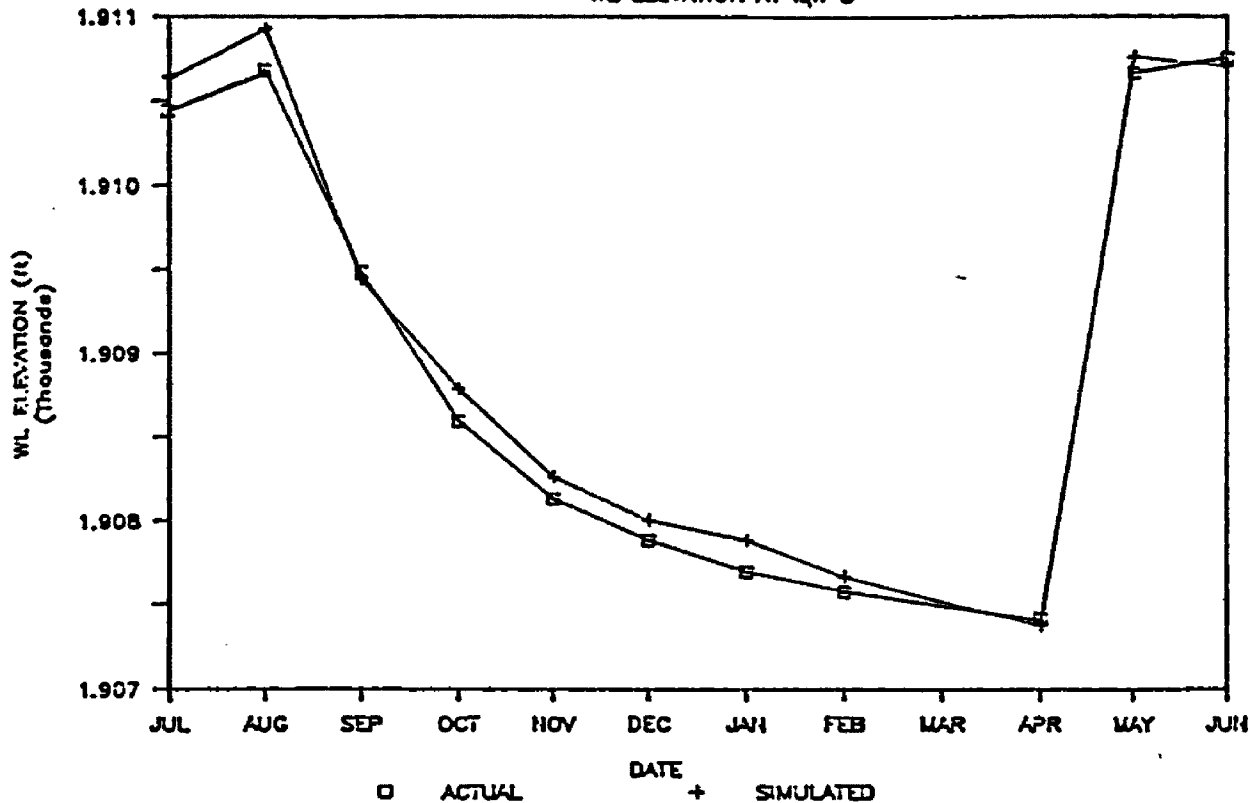


Figure 19. Regression plot of actual and simulated head.

WL ELEVATION AT MW 3



WL ELEVATION AT MW 7s

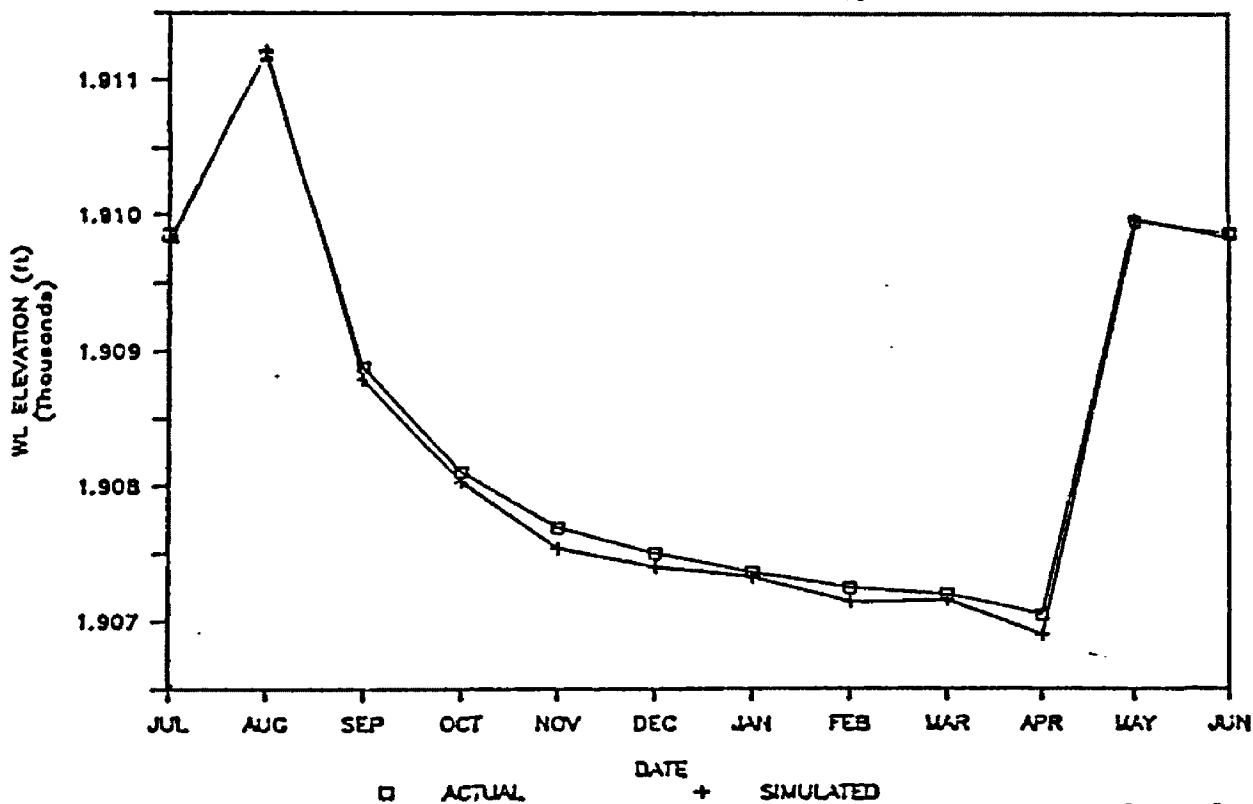


Figure 20. Comparison of actual and simulated water level at MW 3 and MW 7s.

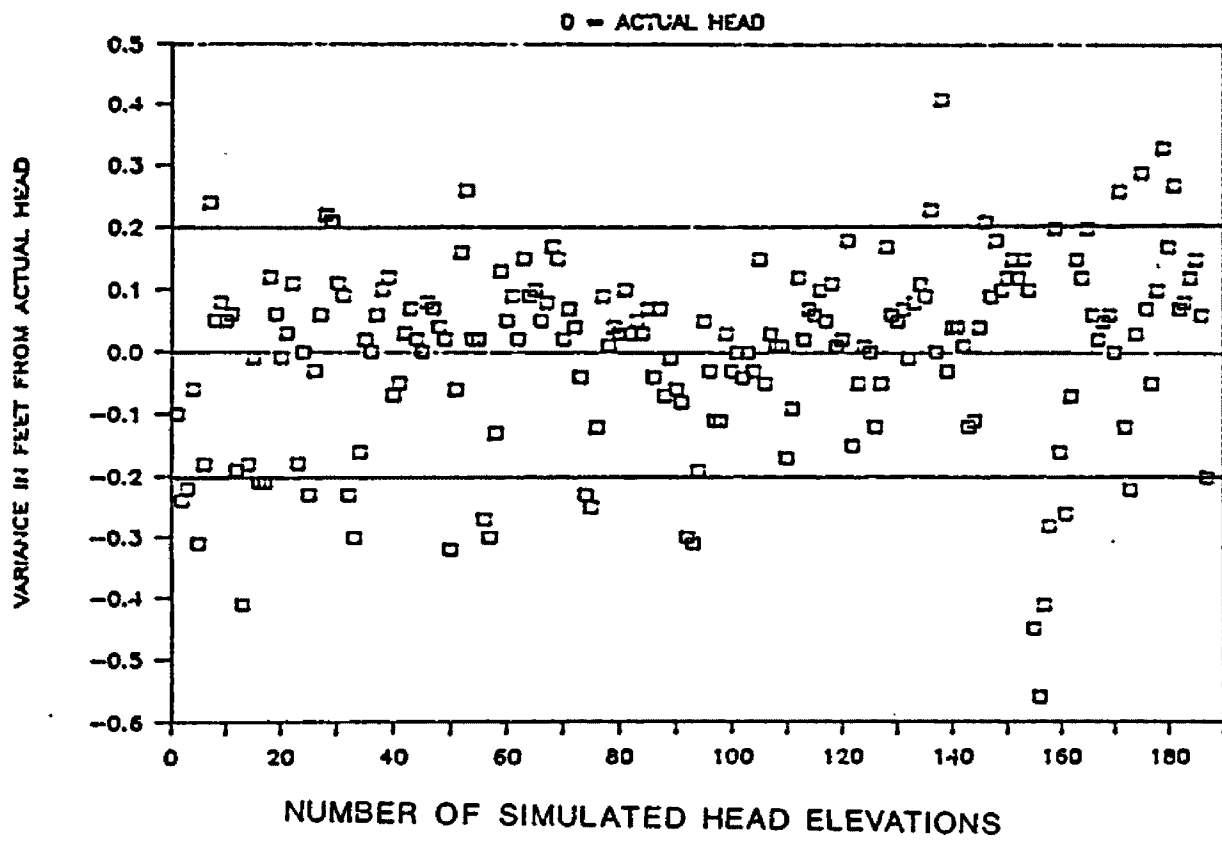


Figure 21. Variance plot of simulated head.

.2 feet of actual head, and greater than 95% of simulated head data are within .3 of actual head data. In addition, the PLASM water balance accounted for 99.9% of all water moving through the model over the duration of the transient simulation.

The modeled hydrogeologic parameters and boundary conditions comprise a relatively simple model in an aquifer that is not homogeneous and isotropic. Modeling head data is not an important factor by itself since monthly groundwater flow is approximate with available water level data. However, modeling different hydrogeologic parameters and boundary conditions provide a better understanding of the hydrogeologic system and were needed to generate flow fields for solute transport modeling.

Solute transport

A preprocessor was used to convert simulated head data from PLASM into twelve groundwater flow velocity fields. They represented flow for each simulated month from July 1987 to June 1988. Chloride transport modeling was initiated in May 1988 since most of 1987 chloride is off site (via the return flow ditch), and a new chloride plume is present at the pit area. The drawback to this method is that chloride was simulated forward in time through July 1988 and then modeled with previous chloride concentration data (August 1987 through April 1988) in the model. This reasoning assumed that the previous year (1987) chloride data (and head distributions)

were 1988 data.

Brine contamination is represented by loading particles into the model, where each particle represents 15 pounds of chloride. Most particles were input to the model during groundwater highs. Particle input was decreased during groundwater declines (Appendix U). In a few months, particle input deviated from the initial approach to increase or decrease chloride concentrations to improve the match of simulated and actual chloride concentrations.

Model input parameters consisted of: porosity (.25), longitudinal dispersivity (20 feet), transverse dispersivity (1 foot), and retardation (= 1; none). Particles were removed from the model if they migrated within a ditch node area since the ditch acts as a groundwater sink during most of the year. Figures 22 and 23 show simulated chloride and actual concentrations for four monitoring well sites. Appendix V contains simulated and actual chloride concentrations at each monitoring well site.

The comparison of actual chloride concentration in monitoring wells with the simulated chloride concentrations were done by comparing the highest chloride level for a specific date at each well site with the simulated value for that location. The chloride concentration at the well sites were assumed to represent the chemistry of the entire saturated thickness of the shallow aquifer (for this reason, only well site identification is presented in Figures 22 and

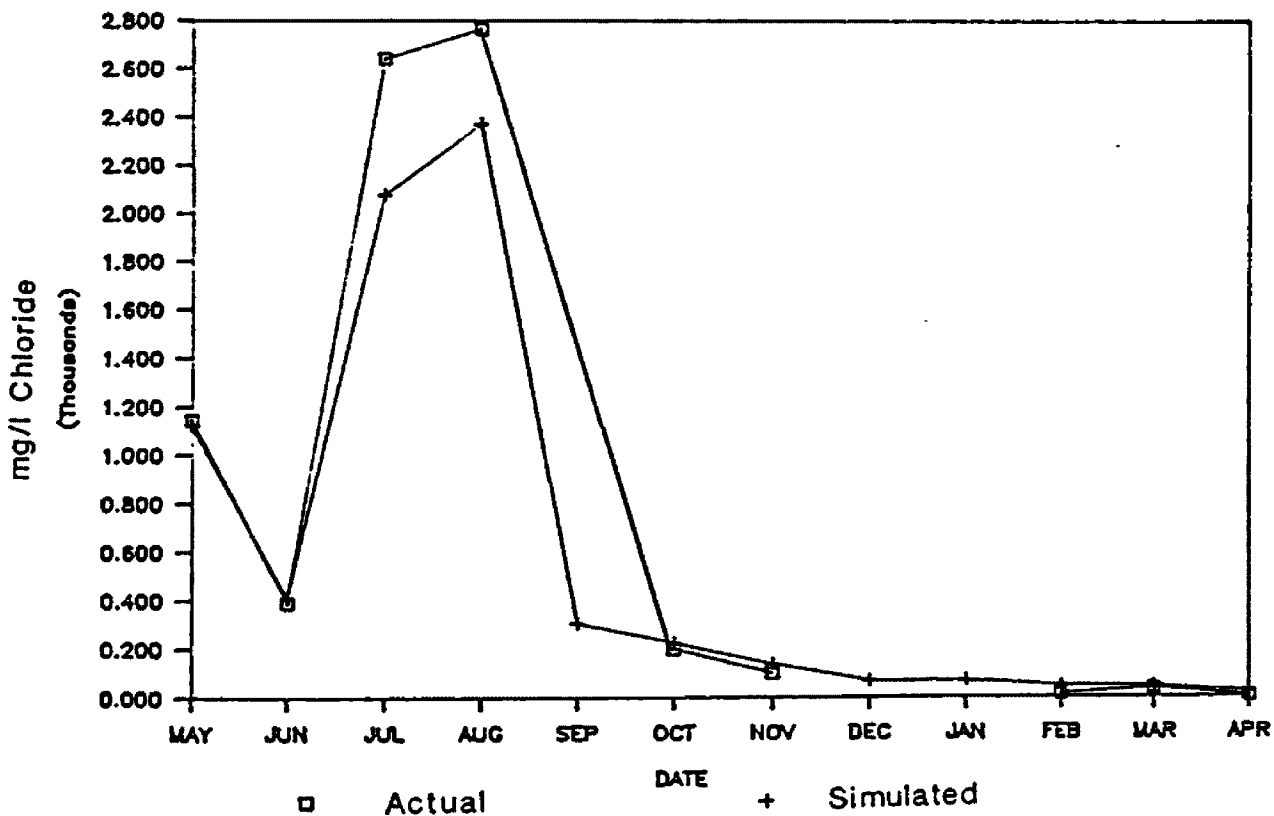
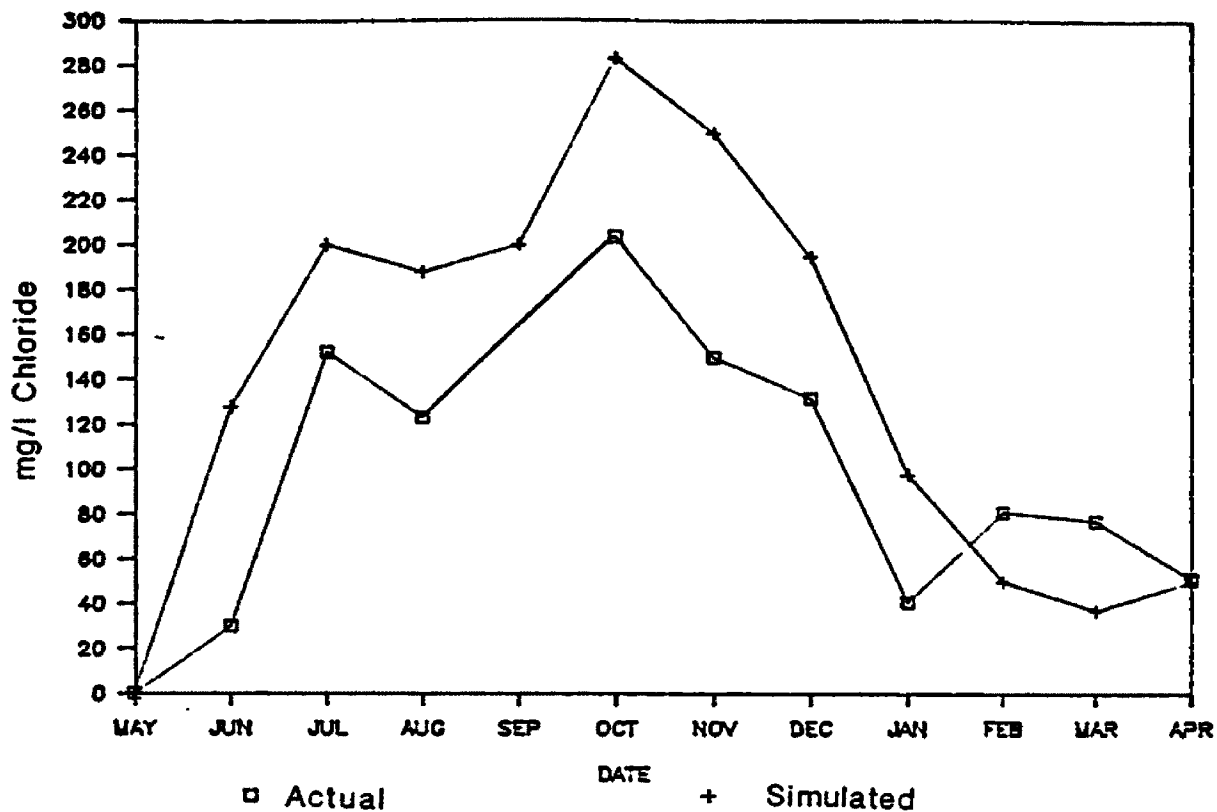


Figure 22. Comparison of actual verses simulated chloride concentrations at well sites MW 7 and MW 3.

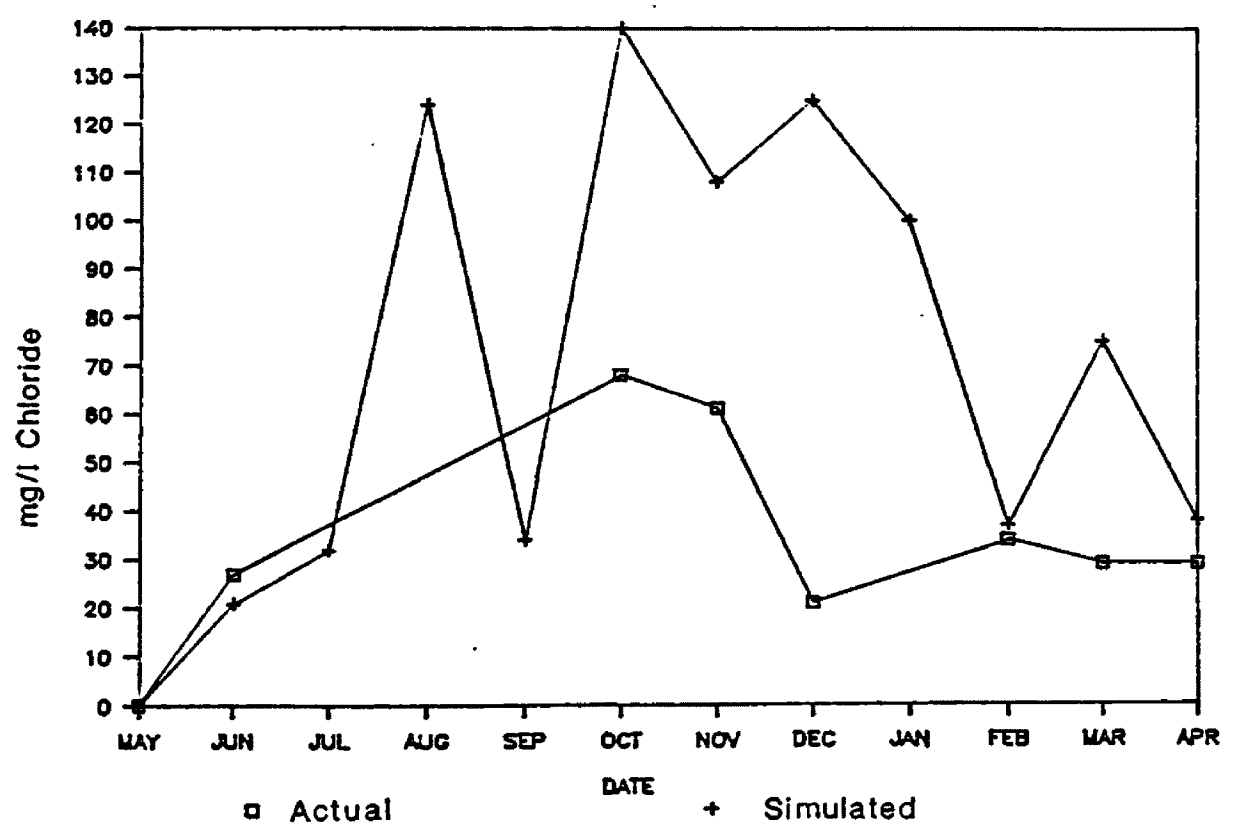
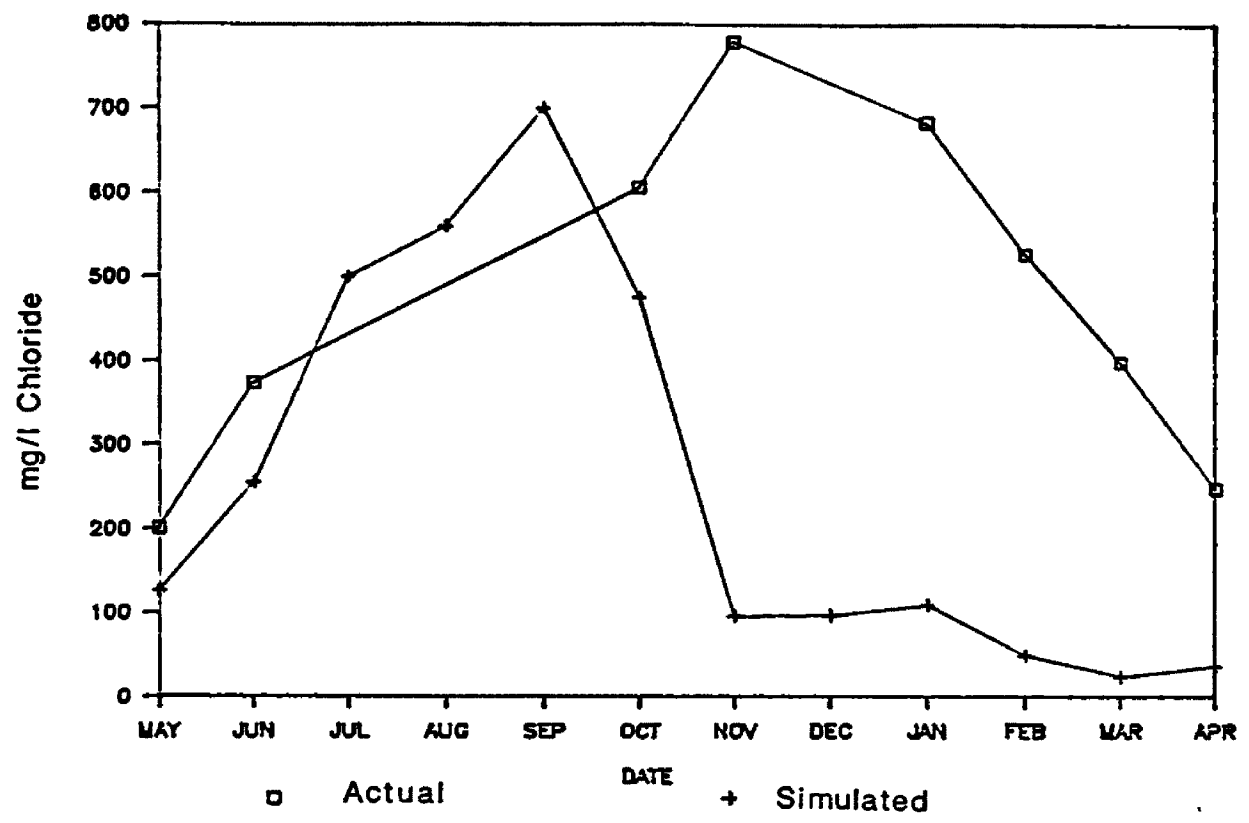


Figure 23. Comparison of actual versus simulated chloride concentrations at well sites MW 16 and MW 19.

23 and no deep or shallow well designations are used to identify well sites).

Figure 22 shows a good correlation between the simulated chloride concentrations at monitoring well sites MW 7 and MW 3. These data indicate simulated chloride concentrations are fairly close to the actual chloride concentrations. However, in Figure 23 the timing of actual versus simulated chloride concentrations at MW 16 and MW 19 is off by two to three months in both diagrams. In addition, the bottom diagram of Figure 23 shows that the simulated chloride concentrations are two to three times greater compared to the actual chloride concentrations. These data indicate the model did an inadequate job simulating chloride concentration through time at these sites. I explain the inaccurate simulated chloride concentrations in a number of ways:

- 1) I modeled my site in two dimensions when in reality it a three dimensional problem. Random-Walk assumes chloride concentration is evenly distributed throughout the entire aquifer thickness which was determined to not be true according to water quality data.
- 2) The flow field was calibrated within .2 to .3 feet error tolerance of actual groundwater elevations which may be insufficient to accurately move particles through the solute transport model.
- 3) Quantab chloride results were sometimes 10% to 50% different from the laboratory chloride concentrations.
- 4) The nodal scale in the model may be too large to accurately simulate head without creating a "blocky" velocity field to move particles in Random-walk.
- 5) Modeling 1988 chloride concentrations with 1987 groundwater flow data may poorly represent the

actual migration and chloride concentrations.

- 6) The flow model may be too simplistic in describing the geologic system. Actual aquifer geology is completed with variable K in three dimensions.
- 7) Solute modeling input parameters may be incorrect.

Annual Chloride load

To calculate annual chloride loading, a cross section extending from MW 5 site, MW 4 site and MW 19s/MW 6 site was used to represent the shallow aquifer. The cross section was divided into blocks or zones vertically and horizontally throughout the saturated zone (Appendix W). Volumes of monthly chloride input from the reserve pit were calculated knowing the volume of groundwater moving through the cross sectional area and chloride concentrations at nested well sites. Monthly chloride data from Quantabs and laboratory chloride analyses were used to determine chloride concentrations. If no data were available, a monthly chloride concentration trend was used to estimate chloride concentrations for a well site.

Monthly chloride loading is in Table 5. Total annual chloride load from July 1987 to June 1988 is calculated to be 3,137 kg/yr from the pit, which is 3.7 times that of background chloride load (852.3 kg/yr chloride). Some error in the load calculations probably occur since about 35% of the chloride concentration data had to be estimated. Other factors which may affect the load accuracy are estimates of groundwater velocity and the area of contamination and

Table 5. Annual chloride load in groundwater (kilograms).

	<u>Pit Input</u>	<u>Background</u>	<u>Simulated</u>
July:	732.45	100.25	2503.9
August:	822.46	94.9	1564.9
September:	450.86	81.43	251.7
October:	242.06	80.56	149.7
November:	113.75	55.44	183.7
December:	70.56	51.9	45.0
January:	56.28	51.9	45.0
February:	40.71	45.28	44.5
March:	47.01	46.55	45.0
April:	44.66	41.58	45.0
May:	210.19	100.25	2245.3
June:	306.01	102.22	898.1
TOTAL:	3137	852.3	8021.9
lbs/yr	6915	1879	17658

chloride analyses/Quantab results.

The total simulated chloride load (Random-Walk simulation) appears to be high compared to the calculated chloride load. However, the model is only two dimensional and must assume the total saturated thickness of the simulated aquifer has the same chloride concentration. Chemical data show that only about half of the shallow aquifer is contaminated with chloride. This required the simulated load be twice as high as the field data in order to reproduce the measured values. The simulated annual chloride is therefore reasonable considering that the lower half of the aquifer is generally at background chloride concentrations and approximately 60% more chloride was input into the simulation compared to the calculated chloride load. The simulated chloride load is greater in the initial and ending months in

comparison to the calculated monthly chloride load. This may be a results of improperly loading chloride in the model.

Elevated chloride concentrations in the ditch water were not detected by chemical analyses. The ditch water varied from approximately 20 to 35 mg/l chloride depending on the ditch flow rate. Chloride data from the up gradient ditch control site did not show any relative increase in chloride concentrations that could be attributed to brine seepage. In some instances the up gradient chloride concentrations were higher by a few mg/l than the down gradient control site where the brine enters the ditch.

Inability to measure the chloride input into the ditch is probably due to the relatively small amount of chloride loaded into the ditch from the reserve pit. This is demonstrated by calculating the chloride concentration which enters the ditch via groundwater (500 mg/l) and the groundwater flux into the ditch (two gpm or 7.57 l/minute). Using these values, 0.057 kg of chloride per minute enter the ditch which produces an increase in chloride concentration of .64 mg/l at an average flow rate of 3.5 cfs. This small amount of chloride loaded into the ditch is not detectable in chemical analyses since the analytical methods used to determine chloride concentration were not precise enough to measure chloride differences of .64 mg/l.

In the winter, when ditch flow is much less, the ditch is loses water to the shallow aquifer. For this reason

chloride is not detected in the winter months.

The pit will continue to load chloride into the shallow aquifer system until the brine contaminants are purged from the pit material. The time necessary to remove available chloride is difficult to determine since controlling factors, including future climate condition, percolation rates and the amount of chloride available in the pit area, are poorly understood or unknown. However, for illustrative purposes an estimate will be attempted.

Parameters used in the attempt include assuming a pit volume (200 feet long by 70 feet wide by 12 feet deep or 61 m by 21 m by 4 m respectively) (Dewey, 1984) and a soil porosity (.25) since open void spaces will not contain chloride. Multiplying these values yields a hypothetical pit volume of 126,000 ft³ or 3,565,800 l of soil. Assuming all chloride in the soil volume loads into the shallow aquifer system, the required time to purge the chloride can be calculated knowing annual chloride load.

Soil moisture analyses at lysimeter L-3 (located in the pit material) indicate the potential amount of removable chloride in the pit material. The largest chloride concentration at L-3 (156,000 mg/l) was used to calculate the volume of removable chloride (where 3,565,800 l x 156,000 mg/l = 5.5626 x 10¹¹ mg or 55,626 Kg of available chloride in the pit material). This value represents the volume of chloride that would take the longest to purge from the pit material.

Assuming the 1987-1988 calculated loading values are the average future chloride loading values, approximately 3,000 kg of chloride are loaded into the shallow aquifer per year. This value is divided into 55,626 kg total available chloride which equals 18.5 years or about 20 years until most of the chloride is removed from the pit from 1988.

The 20 year estimate is probably conservative since it represents a chloride loading pattern similar to the 1987-1988 year for the next 20 years. In reality more time will probably be required to purge all chloride from the pit materials. The "less easily removed" chloride (i.e., small concentrations held up by soil and chemical reactions) will require more leaching action and/or groundwater contact to be removed. This indicates chloride concentrations may be measurable in the shallow groundwater for more than 20 years.

The annual chloride load from the numerous reserve pit sites in the region is probably dependant on the oil-field brine encountered during drilling and the volume and type of drilling additives used during drilling. In addition, not all reserve pit sites may seep brine contaminant into the underlying aquifer. Without knowing the site specific information it is difficult to determine what regional chloride loading impacts will occur in the Yellowstone River Valley as a result of brine seepage from reserve pits.

Reclamation practices and remediation

This and other studies document that reserve pit

reclamation techniques are inadequate. Synthetic or clay liners used to protect the underlying groundwater are intended to inhibit brine seepage. However, when pit liners are ripped or punctured as a result of reclamation, a pathway for brine seepage into the groundwater is created. New cost effective reclamation methods are needed to stop brine from entering the groundwater.

A technique used in some areas is replacement of permanent on site reserve pits with mobil reserve pits. Mobil pits may be more cost effective depending on disposal costs (Reiten, 1988). Drilling fluids can be disposed of by injection wells and solid wastes can be buried at a drilling mud disposal site (properly designed to handle the wastes).

Pit solidification with fly ash has been used for several years in states other than Montana to stop brine seepage from reserve pits. The fly ash causes the mud to be fairly impermeable, limiting brine seepage into the underlying groundwater. This process may be effective, however, it needs to be tested at reserve pits located in the Williston Basin.

Other methods include pumping or evaporating the liquid fraction off the reserve pits and excavating the solid fraction. Brine water can be injected back into saline aquifers and the solid fraction shipped to a central disposal site for burial.

Remediation

Remediation at reserve pit sites with excessive brine

contamination would require one or more of the following:

- 1) The reserve pit and trenches be cleared of mud to eliminate the source.
- 2) A cap over the pit and trenches be installed to limit wetting fronts and slow brine migration (would not affect mud below the water table).
- 3) installation of brine removal pump(s) in severe cases of groundwater contamination.

These are expensive measures and not the only solution to localized brine contamination. Most reserve pit sites are probably similar to the Iverson site and will require little or no remediation. However, highly contaminated sites which threaten vegetation and water supplies need to be cleaned up. Improved pit reclamation techniques could significantly decrease the need to curb brine seepage from reserve pits.

CONCLUSIONS AND RECOMMENDATIONS

Conclusions

- 1) Groundwater quality analyses indicated elevated concentrations of calcium, magnesium, sodium, potassium, chloride, manganese, bicarbonate, sulfate, nitrate, boron, lithium, barium, strontium, titanium, zinc, beryllium and possibly bromide are present in the shallow aquifer at the Iverson site. These contaminants apparently enter the shallow aquifer as a result of wetting fronts moving through the contaminated pit mud and the high water table intersecting the base of the reserve pit.

- 2) Analyses of pit mud show that high concentrations of all constituents detected in the groundwater are in the pit mud in addition to nickel, vanadium, zirconium, aluminum, silver, cadmium, iron, chromium, copper, molybdenum, and lead. These ions were at concentrations a number of orders of magnitude above background levels detected in soil water and groundwater analyses.

- 3) Nitrates and Mn concentrations in the groundwater exceed federal drinking water standards near the pit area in the spring and summer months. Chloride concentrations exceed recommended federal drinking water concentrations year around at the pit area and down gradient of the pit.

- 4) Rising groundwater at the pit area in the spring and summer months, accompanied by greater precipitation, generate high concentrations of brine contaminants in the groundwater. Lowering of the water table in the fall and winter months, accompanied with less precipitation, slows the contaminant loading processes producing lower concentrations at those times.
- 5) High horizontal hydraulic conductivity limits vertical contaminant migration to approximately the upper five to ten feet of the shallow aquifer. Dispersion and dilution appear to lower chloride concentrations by a factor of five at a point 350 feet (115 m) down gradient of the pit.
- 6) The reserve pit will be a groundwater contaminant source for at least 20 years and possibly more.
- 7) Local groundwater flow direction in the spring, summer and fall is partly controlled by the ditch. Groundwater flows is towards the ditch during these months. During winter, groundwater flow is approximately parallel to the ditch. Winter ditch levels are slightly above that of the groundwater, which produces a small groundwater divide below the ditch.

- 8) No water supplies are threatened by brine contamination at this site since groundwater discharge to the ditch removes chloride and other pit contaminants from the shallow aquifer.
- 9) Cumulative groundwater contamination impacts are difficult to measure. However, broad scale regional brine loading impacts to the quality of the Yellowstone River appear to be insignificant. Local brine contamination from oil well sites could have serious local effects on water supplies and vegetation depending on the geology, groundwater flow direction and contaminant loading rate. The potential of these effects should be considered when siting and reclaiming reserve pits.
- 10) EM is a valuable tool to delineate brine contamination and plume migration at oil well sites. EM equipment can be used to survey a site quickly and easily to determine the extent of brine contamination. Resistivity delineates brine contamination as accurately as EM but it is cumbersome and time-consuming to use.
- 11) Groundwater flow and solute transport modeling are important tools to improve the understanding and characterization of the hydrogeologic system, aquifer

parameters, and contaminant migration.

- 12) Flow modeling accurately simulated flow direction, head and groundwater velocities.
- 13) The accuracy of simulated chloride concentrations were limited by the simplicity of the groundwater flow model, Random-Walk limitations of dispersion and dilution, boundary conditions and model inputs. Results show chloride concentrations were accurately simulated at some locations and inadequately simulated at other locations.
- 14) Current reclamation practices in Montana are inadequate to prevent brine contamination in shallow alluvial aquifers.

Recommendations

Individuals pursuing research involving reserve pit contamination should consider the following topics:

- 1) Determine whether organic contamination is a potential threat to groundwater quality at oil well sites drilled with oil based mud or mud containing diesel fuel additives. Volatile, semi-volatile and long chain organic compounds are potential contaminants not examined in current reserve pit literature.

- 2) Perform EM work needed to located contaminated oil well sites in eastern Montana, and correlate chloride concentrations in the groundwater with EM conductance values.

- 3) Determine cost effective measures that can be implemented to replace current reclamation techniques to stop or reduce brine contamination.

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APPENDIX A
FUNCTION AND GENERAL PURPOSE OF DRILLING FLUIDS ADDITIVES

Function and general purpose of drilling fluid additives (adapted from Murphy and Kewhew, 1985)

<u>Function</u>	<u>General Purpose</u>	<u>Common Additives</u>
Weighting Material	Control formation pressure, check caving, facilitate pulling dry pipe, and well completion operations	Barite, lead compounds, iron oxides
Viscosifier	Viscosity builders for fluids, for a high viscosity-solids relationship	Bentonite, attapulgite clays, all colloids, fibrous asbestos
Thinner Dispersant	Modify relationship between the viscosity and percentage of solids, vary gel strength, deflocculant	Tannins (Quebracho), polyphosphates, lignitic materials
Filtrate Reducer	Cut the loss of the drilling fluid's liquid phase into the formation	Bentonite clays, sodium carboxymethyl cellulose (CMC), pregelatinized starch, various lignosulfonates
Lost Circulation Material	Primary function is to plug the zone of loss	Walnut shells, shredded cellophane flakes, thixotropic cement, shredded cane fiber, pig hair, chicken feathers, etc.
Alkalinity, pH Control	Control the degree of acidity or alkalinity of a fluid	Lime, caustic soda, bicarbonate of soda
Emulsifier	Create a heterogeneous mixture of two liquids	lignosulfonates, mud detergent, petroleum sulfonate
Surfactant	Used to the degree of emulsification, aggregation, dispersion, interfacial tension, foaming, and defoaming (surface active agent)	Include additives used under emulsifier foamers, defoamers, and flocculators
Corrosion Inhibitor	Materials attempt to decrease the presence of such corrosive compounds as oxygen, carbon dioxide, and hydrogen sulfide	Copper carbonate, sodium chromate, chromate-zinc solutions, chrome lignosulfonates, organic acids and amine polymers, sodium arsenite
Defoamer	Reduce foaming action especially in salt-water-based muds	Long chain alcohols, silicones, sulfonated
Foamer	Surfactants which foam in the presence of water and thus permit air or gas drilling in formations producing water	Organic sodium and sulfonates, alkyl benzene sulfonates
Flocculants	Used commonly for increases in gel strength	Salt, hydrated lime, gypsum, sodium tetraphosphates
Bactericides	Reduce bacteria count	Starch preservative, paraformaldehyde, caustic soda, lime, sodium pentachlorophenate
Lubricants	Reduce torque and increase horsepower at the bit by reducing the coefficient of friction	Graphite powder, soaps, certain oils
Calcium Remover	Prevent and overcome the contamination effects of anhydrite and gypsum	Caustic soda (NaOH), soda ash, bicarbonate of soda, barium carbonate
Shale Control Inhibitors	Used to control caving by swelling or hydrates disintegration	Gypsum, sodium silicate, calcium lignosulfonates, lime, salt

APPENDIX B
DISCUSSION ON THE REGIONAL HYDROGEOLOGY

The site is located on the western flank of the Williston Basin (Exhibit 1) in Richland County, Montana which is a deep structural depression containing 14,000 to 15,000 feet (4500 m) of sedimentary rock overlying a gneissic precambrian basement (Carlson, 1985).

As of 1982 oil removed from the Williston basin in Montana had totaled 21,923,760 barrels from 1,360 wells (Dewey, 1984). Major oil producing zones 9,000 to 13,000 feet (2700 to 4000 m) below ground surface include the Paleozoic Mission Canyon Limestone and Red River formation (Prichard and Landis, 1975). The Spearfish Evaporite sequence is above the oil producing zones and marks the Permian boundary (Carlson, 1985).

The Regional surficial geology encompasses the Tertiary Fort Union Formation to the Quaternary alluvial and fluvial deposits. Exhibit 2 shows a geologic cross section of the area and Exhibit 3 describes the Mesozoic through the Quaternary stratigraphy. The Fort Union formation has four members, however, only the Tongue River member is significant to this study. The Tongue River Member is a light grayish to brown layered sand, silt, and clay deposit with .25 to five feet thick coal beds and clinker zones. Gypsum layers and iron nodules are found in many of the silt and clay layers. Most beds are massive with local crossbedding, and weathered to a yellowish buff color. Cementation is weak in most rock layers making the Tongue River Member soft. The depositional environment was probably similar to a terrestrial flood plain, marsh, or swamp setting (Prichard and Landis, 1975). The Tongue River Member generally yields approximately 10 gpm of water from

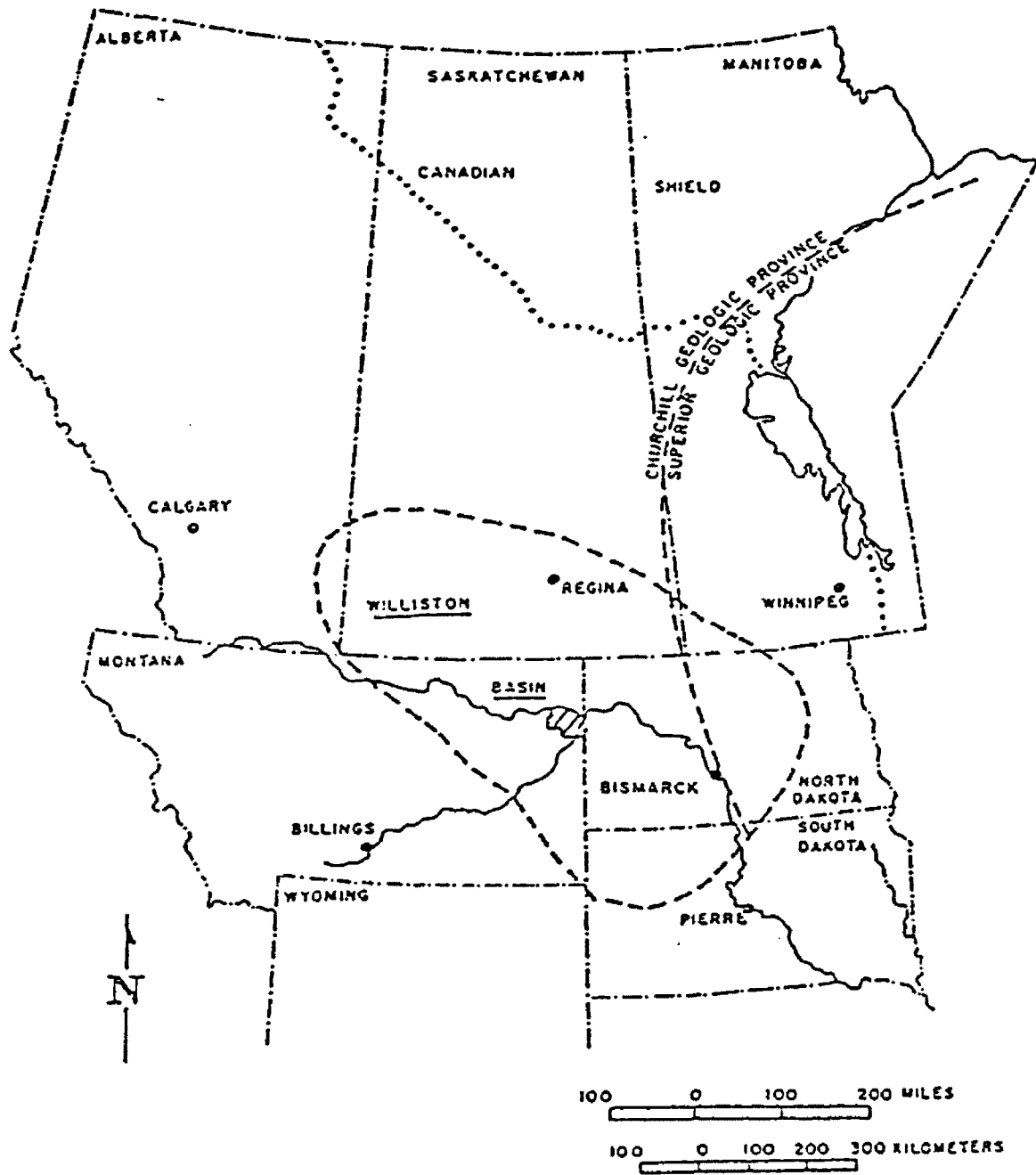


Exhibit 1. Location of Richland County Montana, Montana in the Williston basin shown by hatched lines (adapted from Dewey, 1984).

EAST

WEST

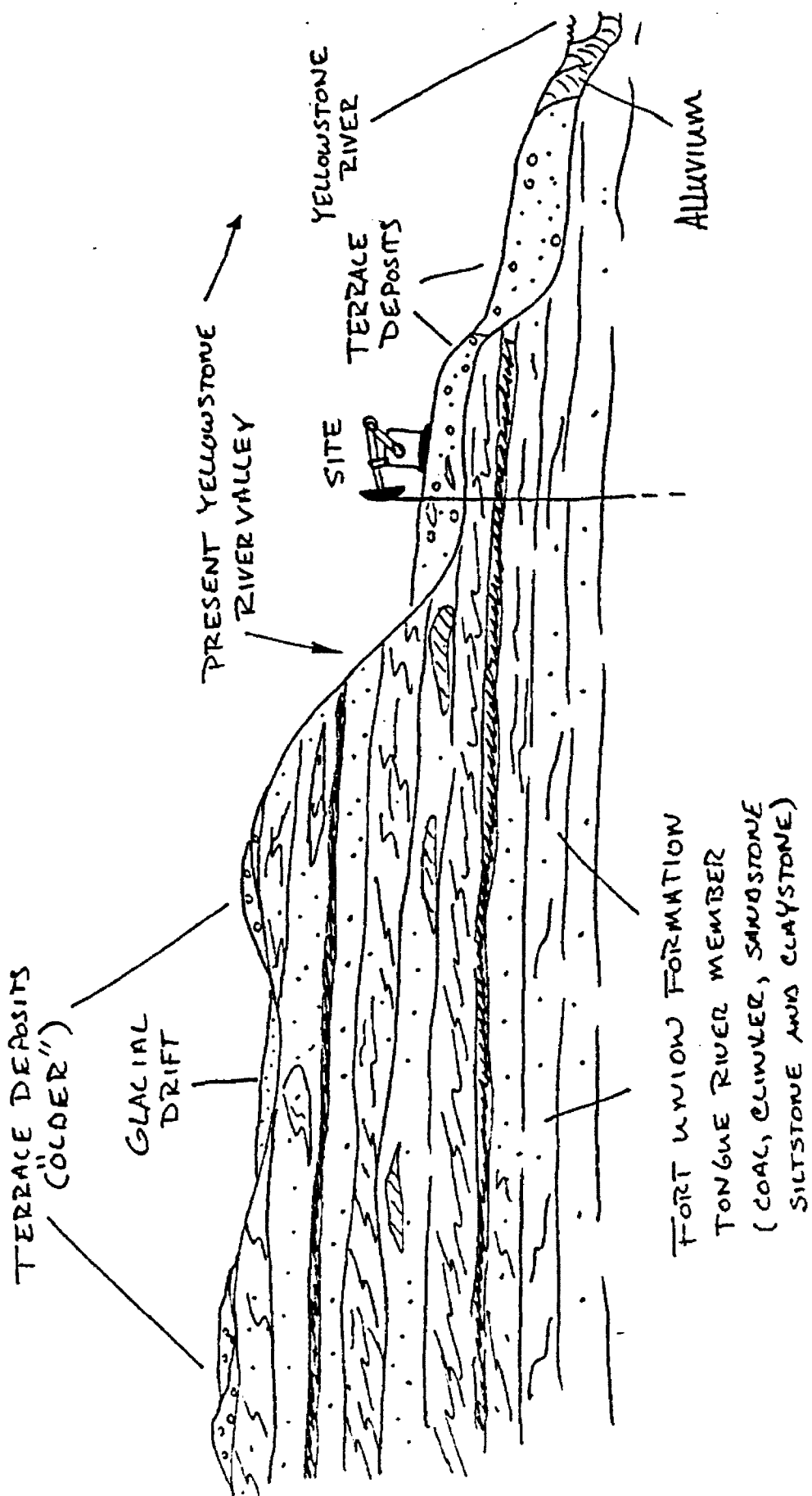


Exhibit 2. Geologic cross section of the study area (diagrammatic).

Stratigraphic Unit	Section	Correlation Unit	Thickness (feet)	General description	Water-bearing characteristics	
Quaternary	Nashua and Pleistocene	Alluvium	0-130	Mostly unconsolidated sand, silt and clay with local lenses of gravel. Some well-sorted gravel interbedded with finer material is common in basal alluvial deposits along the Yellowstone and Missouri Rivers. Some smaller streams gravel consists of silty sandstone fragments and broken pebbles of quartz, feldspar and mica. Thickness of the unit is as much as 130 feet along the Yellowstone River and 40 feet along smaller streams.	Gravelly sands along major streams are reported to yield as much as 1 MGD per acre to large diameter wells. Along smaller streams with less heterogeneity, some yields of 25 gal per acre have been reported. Yields commonly are 30 gal per acre to steel and concrete wells.	
		Terrace deposits	0-100	Mostly gravel and sand with some silt and clay. Well rounded to sub-angular pebbles and cobbles of quartzite and some siliceous basaltic of quartzite, granite, and igneous rocks are common. Deposits are largely disarticulated along the channel of the Yellowstone River. Includes the Grand Teton and Carwright Gravels.	Terraces are usually lower permeability and isolated units having limited water. Yields to concrete and steel wells are as high as 20 gal per acre from larger diameters at least six miles along major streams.	
		Clayey drift	0-70	Medium clayey fine sandstone containing scattered pebbles and cobbles and occasional fine boulders. Locally silty with a gravelly sandstone upper part. Contains local lenses of gravel. Gravelly sandstone contains pebbles and cobbles of quartzite, granite, and some siliceous basaltic other than granite. Locally silt clay but contains some local pebbles in upper part. Consists of gravel outside and around deposits.	Yields as a source of water in the northern part of the study area, generally within the area covered by continental glaciation. Yields to wells are generally less than 20 gal per acre. Larger yields may be available from more extensive gravel lenses.	
Cambrian	Yellowstone	Pliocene (?) or Miocene (?)	Fluvial formation	0-100	Yellow to tan-gray fluvial gravel, sand, and silt with local pebbles and cobbles. Gravelly sandstone of well-sorted sand, silt, and gravel. Pebbles average about 1 inch in diameter, but boulders as much as 12 inches in diameter are included. Locally includes cross-bedded, hard sandstone, and calcite-cemented conglomerate. Well exposed to be compared to the Fluvial Formation may be compared to the slightly older Niangua (?) or Niangua (?) Grand Canyon.	A possible source of water to areas of permeability in the northern Yellowstone River study area. Yields to wells of as much as 10 gal per acre have been reported. Yields to steel and concrete wells are less than 20 gal per acre. Where saturated, fractured clays may yield as much as 20 gal per acre.
			Upper Sandstone Member	0-1,200	Light-gray to brownish-gray fine to medium-grained sandstone to siltstone and shale cross-bedded, laminated sandstone and siltstone. Commonly contains light yellow to buff. Contains light-gray to tan-gray shaly siltstone and shale. Some of the sandstone is white and coal-bone. Shifting of coal during deposition has caused irregular sandstone and shale to form into a lenticular pattern.	Sandstone and shale beds contain the best water-bearing unit. Siltstone and shale are fluid impermeable substrates of water. Yields to wells of as much as 10 gal per acre have been reported. Yields to steel and concrete wells are less than 20 gal per acre. Where saturated, fractured clays may yield as much as 20 gal per acre.
			Lower Sandstone Member	0-100	Franklinville dark shale interbedded with light-gray and brown to black carbonaceous shale, siltstone, and locally thin sand. Some sandstone contains calcite and calcification. Some shaly sandstone contains some calcite. Shale to light-gray argillaceous cross-bedded and lenticular sandstone occur locally.	A limited source of water to the study area. Relative permeability is low and yields to wells are as much as 10 gal per acre have been reported. Yields to steel and concrete wells are less than 20 gal per acre. The permeable nature of these yields to the lenticular sandstone that occur locally within the unit.
			Tuffaceous Sandstone	0-100	Interbedded sandstone and light-gray shale, light-gray fine-grained sandstone and siltstone, and thin red sandstone and shale. A resistant sandstone commonly forms a firm at the top of the unit. Locally, sandstone and siltstone weather yellow to brown.	Thin-grained sandstone and shale have very small quantities of water for steel and concrete wells. Yields to wells of as much as 10 gal per acre have been reported but generally are less than 15 gal per acre.
			Well Creek Formation	0-400	Gray to yellowish-gray silty clayey sandstone and siltstone and locally shaly sandstone. Locally a calcareous gray to tan fine to medium-grained silty sandstone containing thin coal beds, sandstone, shale, and thin coal beds. Some of the sandstone is calcareous and contains thin coal beds.	Upper part of Well Creek is limited as a water source to the study area. Well yields of as much as 10 gal per acre have been reported but generally average about 5 gal per acre.
			Fox Hills Sandstone	0-400	Upper part consists of very light gray fine to medium-grained sandstone. Lower part contains gray to brownish-gray fine-grained sandstone and siltstone interbedded with gray sand shale and siltstone.	Lower part of Well Creek and Fox Hills Sandstone is considered to be no water to the study area. A significant amount of water for steel and concrete wells in the study area. Vertical flows of as much as 10 gal per acre have been reported and the total yields of 100 gal per acre have been reported from large-diameter wells.
Precambrian	Upper Cambrian	Basal Group	Beaverhead Shale	400-1,100	Gray to black carbon shale and shale siltstone thin sand of siltstone, silt sandstone, and bentonite bedded shaly.	Very low permeability. Generally does not yield water to wells in the study area.
			Yellow River Formation	25-150	Buff to white massive fine to medium-grained sandstone. Contains interbeds of gray to black, tan, carbonaceous shale and silt shale and some light gray sand.	Because of the large depth of this unit and the availability of large quantities of water at shallower depths, this unit does not constitute an important source of water in the study area. Yields to wells of as much as 20 gal per acre have been reported.

Exhibit 3. Stratigraphic column for Richland County, Montana (from Slagle, 1981)

wells, most of which is derived from coal beds. Many wells are completed in the Tongue River Member since the water tends to be softer than other near surface water supplies and wells are generally not deeper than 75 feet.

The Yellowstone River deposited terrace gravel upon the Fort Union formation. These gravel were previously named the "Flaxville Gravel" which has now been abandoned. The oldest terrace deposits as reported by Howard (1960) are Miocene or Pliocene and may have a surface younger than Pliocene. The gravel cap the Fort Union highs and are composed of coarse gravel and cobbles in a coarse to fine sand matrix, and in some instances, the deposits are weakly cemented with calcite.

The Yellowstone River cut into the oldest terrace deposits and formed the Missouri Plateau peneplain, on which a younger terrace gravel were deposited (previously named the Cartwright Gravel). These gravel deposits are early Pleistocene age laid down by the ancestral Missouri and Yellowstone Rivers (Howard, 1960) and found on topographically high Fort Union banks as erosional remnants of a once extensive gravel blanket. These terrace deposits ("Cartwright Gravel") are lithologically similar to the older terrace deposits ("Flaxville Gravel"), and may in fact be reworked terrace deposits (Alden, 1932). Both of these deposits ("Cartwright and Flaxville") generally do not yield more than 10 gpm and the quality of the water is often poor due to a high TDS and hardness.

Another period of erosion and aggradation period deposited a younger terrace gravel sequence (previously named the Crane Creek

Gravel). These gravel exhibit a similar lithology and texture as the two older terrace deposits and hence, thought to be in part derived from these deposits (Prichard and Landis, 1975). The major difference between the two older terrace gravel and the younger terrace gravel ("Crane Creek Gravel") is that the younger it is constrained to the present Yellowstone drainage, and is not preserved on plateaus and broad interstream uplands of the ancestral Missouri and Yellowstone Rivers. Howard (1960) correlates these gravel as pre-Wisconsin Pleistocene or Yarmonth Interglaciation. These younger terrace deposits produce up to 1,000 gpm in the major river valleys and are recharged by irrigation and channel seepage in addition to precipitation.

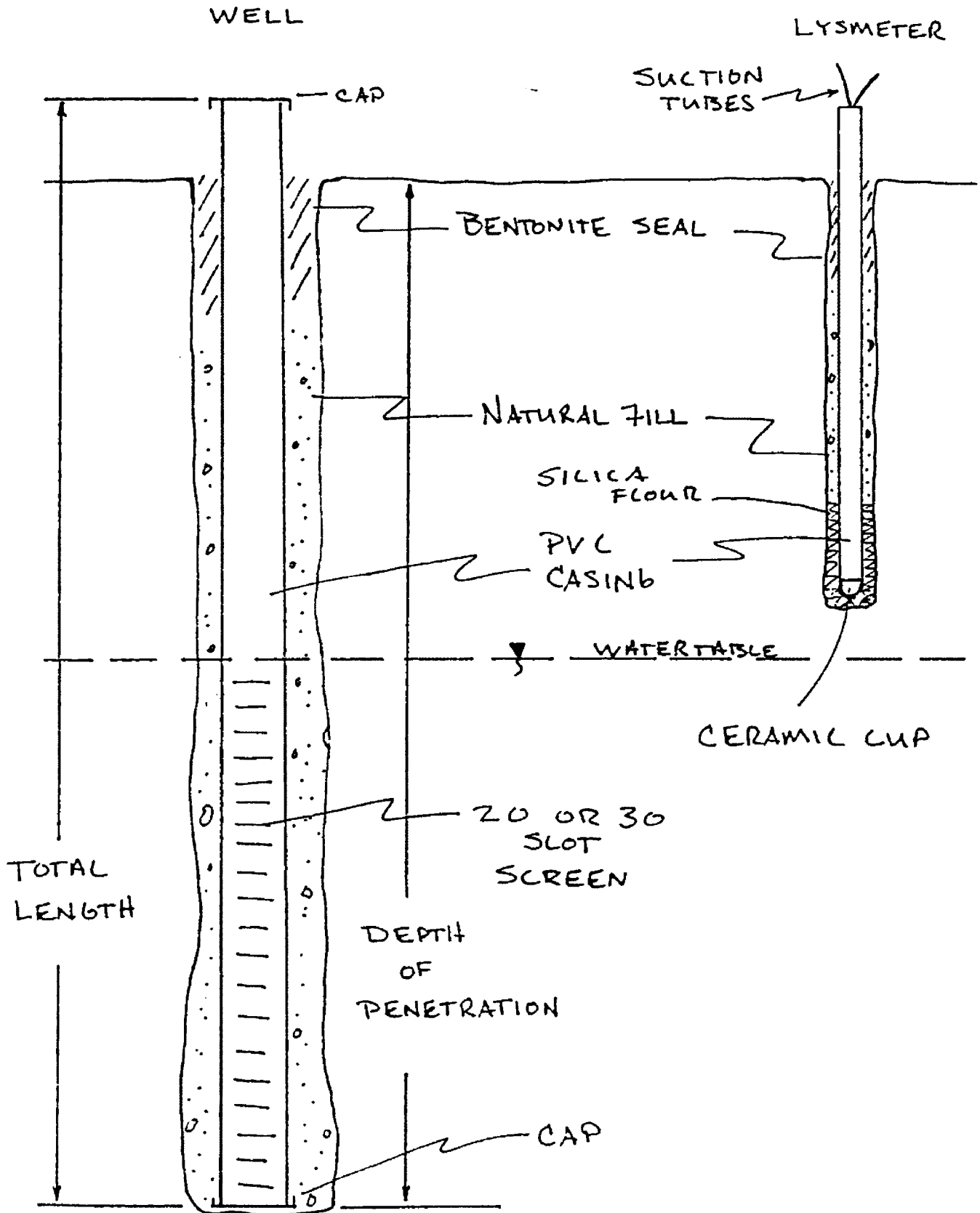
These younger terrace deposits ("Crane Creek Gravel") are infrequently covered by Glacial till, but do show evidence of continental glaciation after deposition. Glacial drift is early Wisconsin (?) according to Howard (1960) and Iowan or Illinoian (?) according to Alden (1932). Glacial deposits include ground moraines, stratified drift, till, melt-water and diversion-channel deposits, eskers, and kames (Prichard and Landis, 1975). The exact number of ice advances and glacial interims is disputed and not discussed in this text (see Alden, 1932; Carlson, 1985; Howard, 1960; and Prichard and Landis, 1975). Glacial deposits are reported to produce 20 gpm of water from wells but are seldom utilized.

Alluvium in the Yellowstone River Valley is mapped as two older flood plain terraces and alluvium deposited within the river channel. Alluvial deposits are predominately Holocene age

deposited after an erosional period, which topographically has changed little since deposition of the older alluvium (Howard, 1960). Exhibit 3 is a stratigraphic column of units discussed above which also reviews water-yielding characteristics of each geologic unit. Alluvial deposits generally yield 10 gpm.

APPENDIX C
WELL CONSTRUCTION

TYPICAL LYSIMETER AND WELL CONSTRUCTION



WELL CONSTRUCTION INFORMATION

WELL #	ELEVATION	CASING	SCREEN SLOT/LENGTH	TOTAL LENGTH	DEPTH OF PENETRATION
MW 1	1922.1'	2"PVC	#20/3.0'	21.75'	20.0'
MW 2	1920.14'	2"PVC	#20/5.0'	20.0'	18.0'
MW 3	1920.14'	2"PVC	#20/5.0'	20.0'	18.17'
MW 3s	1920.89'	2"PVC	#30/2.25'	14.0'	11.4'
MW 4s	1921.09'	2"PVC	#30/2.0'	14.0'	12.33'
MW 4a	1919.74'	2"PVC	#30/2.0'	16.0'	14.17'
MW 4d	1922.24'	2"PVC	#30/2.0'	22.0'	19.75'
MW 5	1919.59'	2"PVC	#30/5.0'	20.0'	18.08'
MW 5s	1919.34'	2"PVC	#20/3.0'	13.5'	11.5'
MW 6	1921.14'	2"PVC	#30/5.0'	22.0'	19.0'
MW 7s	1919.84'	2"PVC	#30/1.5'	15.3'	13.8'
MW 7d	1919.69'	2"PVC	#30/2.0'	24.0'	22.17'
MW 8	1915.72'	2"PVC	#20/3.0'	22.0'	20.5' (?)
MW 9	1920.48'	2"PVC	#30/2.0'	22.2'	19.62'
MW 10	1916.67'	2"PVC	#30/3.0'	23.0'	20.17'
MW 11	1919.74'	2"PVC	#30/3.0'	23.0'	21.75'
MW 12	1917.77'	2"PVC	#20/3.0'	22.0'	20.0'
MW 13	1919.09'	2"PVC	#20/3.0'	22.0'	19.5'
MW 14	1921.54'	2"PVC	#20/5.0'	23.0'	21.33'
MW 15d	1919.84'	4.5"PVC	#20/14.0'	21.1'	18.6'
MW 16s	1919.87'	2"PVC	#20/6.0'	14.0'	12.0'
MW 17s	1918.18'	2"PVC	#20/3.0'	15.0'	12.6'
MW 18s	1917.75'	2"PVC	#20/3.0'	13.0'	11.0'
MW 19s	1920.74'	2"PVC	#20/6.0'	17.0'	14.8'
MW 20s	1917.96'	2"PVC	#20/3.5'	13.5'	12.8'
MW 21e	1920.64'	2"PVC	#20/3.0'	15.0'	12.4'

LYSIMETER CONSTRUCTION INFORMATION

LYSIM. #	CASING	TOTAL DEPTH	DEPTH OF PENETRATION	
L2	1.5"PVC	5.17'	3.59'	
L3	1.5"PVC	6.83'	6.41'	
L3s	1.5"PVC	4.25'	2.92'	
L3SD	(NONFUNCTIONAL)			
L3SS	(NONFUNCTIONAL)			
L8	1.5"PVC	5.42'	3.84'	(DESTROYED 10/87)

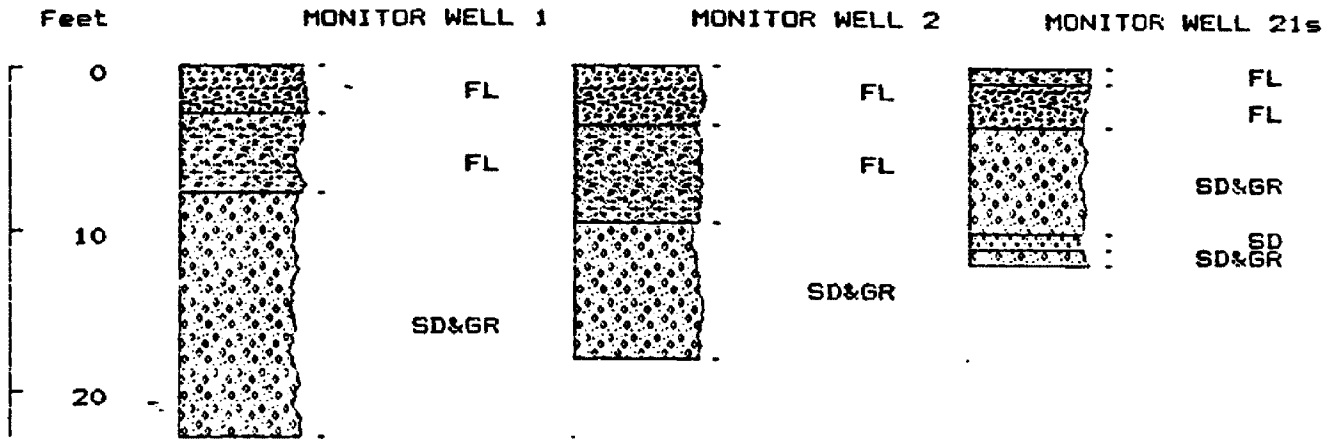
APPENDIX D
MONITOR WELL LOGS, SIEVE ANALYSES AND SITE GEOLOGY

LITHOLOGY KEY

SO	SO - Blackish tan/dark brown soil, fine to coarse sands with 5 to 15% silt and clay, and 5 to 20% gravel, moist, moderate to poorly sorted, rounded and subangular.
SO	SO - Blackish tan to green soil, fine grained sand with 25 to 35% silt and clay, moist, moderately well sorted, subrounded to subangular, a few gravel bits and clay lenses.
FL	FL - Blackish red fill soil and debris, coarse to fine sand, gravel and silts with 5 to 10% clay (?), v. poorly sorted, subrounded gravel and cobbles, 5% coarse clinker bits, trace wood and plastic debris.
FL	FL - Blackish brown and grayish fill soil, medium to v. fine sand with 10% coarse grains and 20 to 30% silts and clays, 5 to 15% gravel, poorly sorted, subrounded to subangular, 5% clinker bits.
SD	SD - Tan brown to green clean sand, medium to fine grained with 5 to 5% gravel, fairly well to well sorted, subrounded to subangular, Qtz rich with 5 to 10% dolomite and clinker, 2% mica, and trace coal bits.
SD	SD - Tan brown fairly clean sand, medium grained with 10 to 20% pea gravel, fine and coarse sand, moderately sorted, subrounded to subangular sand and subrounded gravel, Qtz rich, 5% clinker, dolomite, agates, and coal bits.
SD&GR	SD&GR - Brownish sand and gravel, fine to coarse sand with .5 to 3 inch gravel, 5% cobbles, 2 to 10% silt and clay, v. poorly sorted, subangular sand and subrounded gravel, Qtz rich with siltstone, granite, gneiss, dolomite, sandstone, and acate clasts. coagon aquifer material for site.
GR	GR - Brownish clean gravel, 1 to 2 inch gravel with 10 to 20% smaller and larger, 5 to 10% cobbles, 2 to 5% fines, moderately to poorly sorted, subrounded to subangular, Qtz rich with siltstone, granite, gneiss, dolomite, sandstone, and acate clasts. channel cut deposit (?).
CL	CL - Dark green to gray pit mud, silty/clayey fine sand with a few clinker bits and pea gravel, 30% silt and 20% clay, moderately sorted, rounded to subrounded clasts, moist, oily saell (diesel?).
SL&CL	SL&CL - Gray to dark green clayey silt, trace gravel, moderately sorted, rounded to subrounded clasts, v. moist, oily saell near pit area-possibly a trench dug from the reserve pit(?), more tan colored away from the pit.
CL/SD	CL/SD - Brown interlayered clay and sand with 20% gravel, v. poorly sorted sands and fairly sorted clays, subrounded, Qtz rich, trace clinker bits.
SL/SD	SL/SD - Brown interlayered silt and sand with 15 to 20% gravel, poorly sorted sands and fairly sorted silts, subrounded, Qtz rich.
SL/GR	SL/GR - Gray to Brown clayey-silty sand with 5 to 10% pea gravel, 50% silt and clay, medium to coarse sand, moderate to poorly sorted, moist, till(?).
FU	FU - Fort Union bedrock, thick layers of gray silty clay and black coal beds, 20% fine sandy lenses and trace pea gravel in silty clay, fairly well sorted, rounded to subangular, v. moist to wet, v. dense "siltstone".

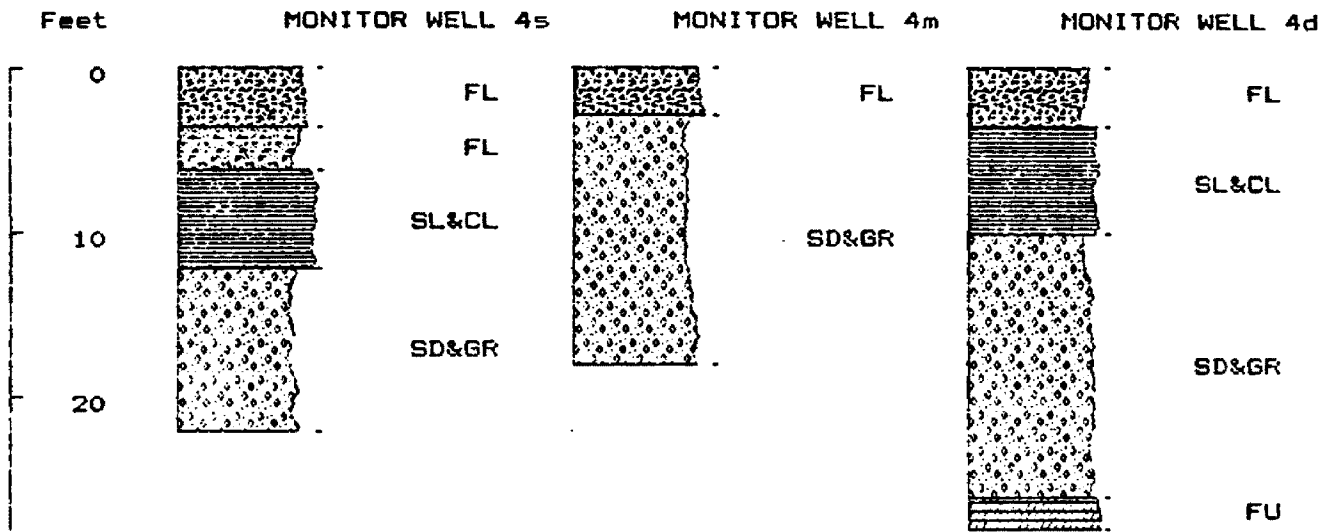
ETNA TRILOBITES STRATCOL

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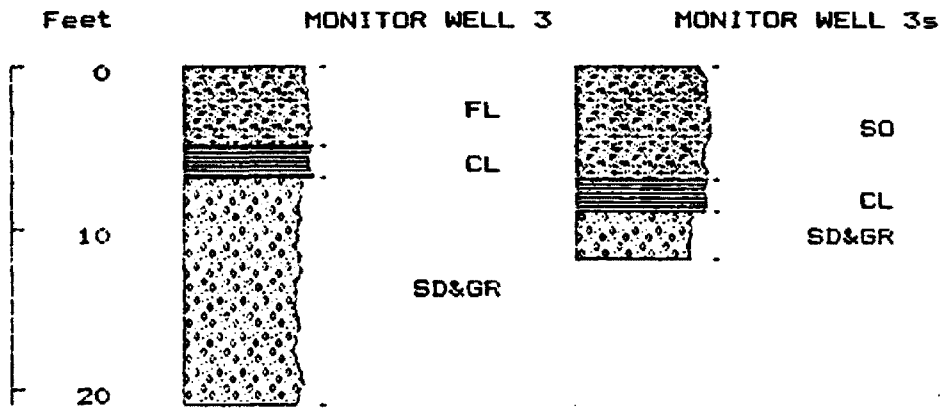
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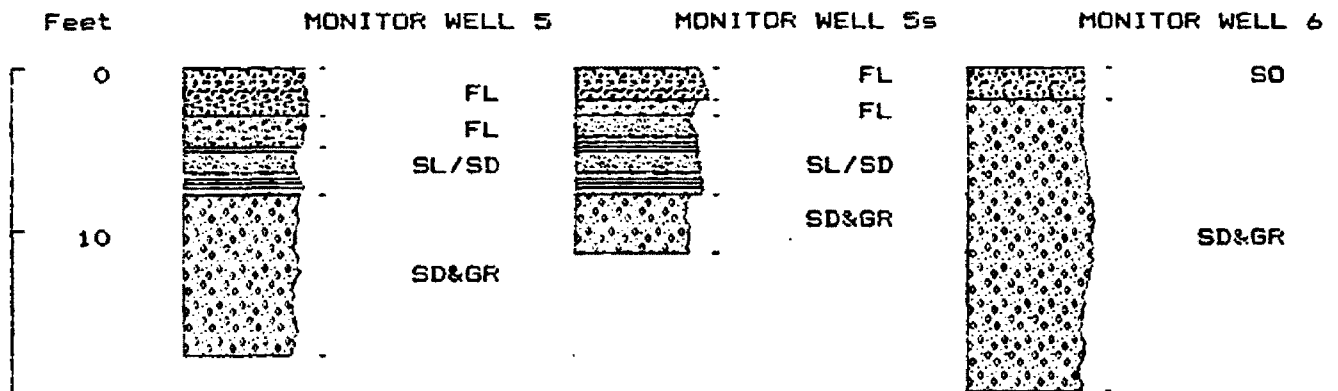
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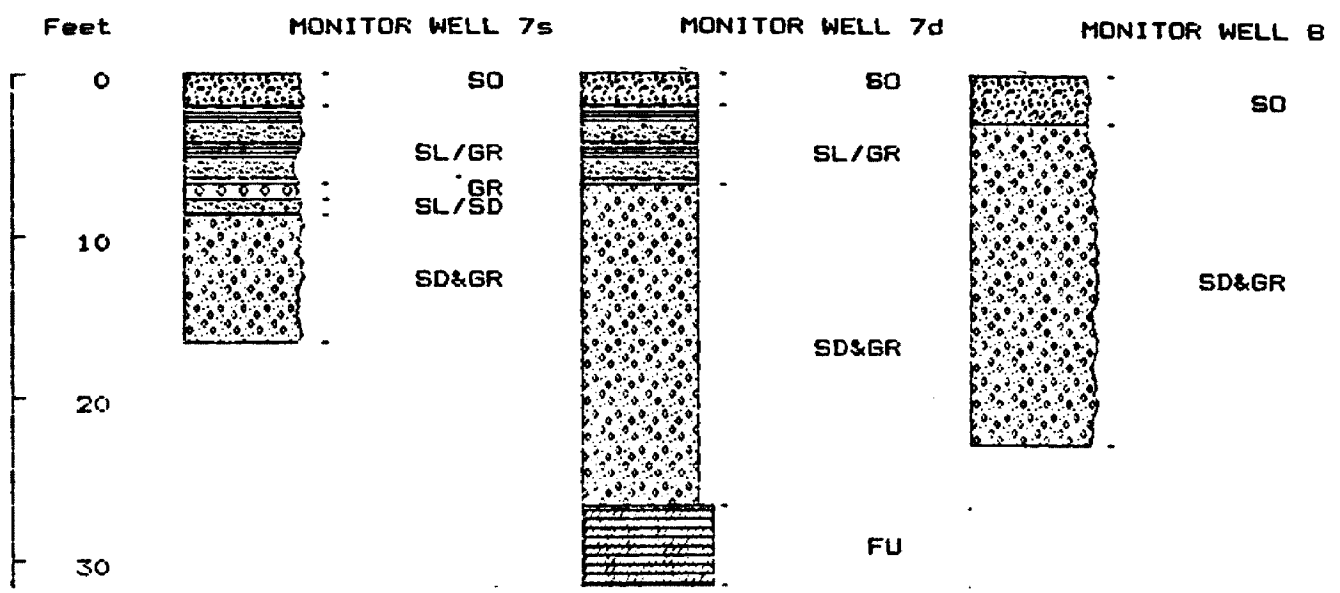
ETHAN Trilobites STRATCOL

Scale 1: 120



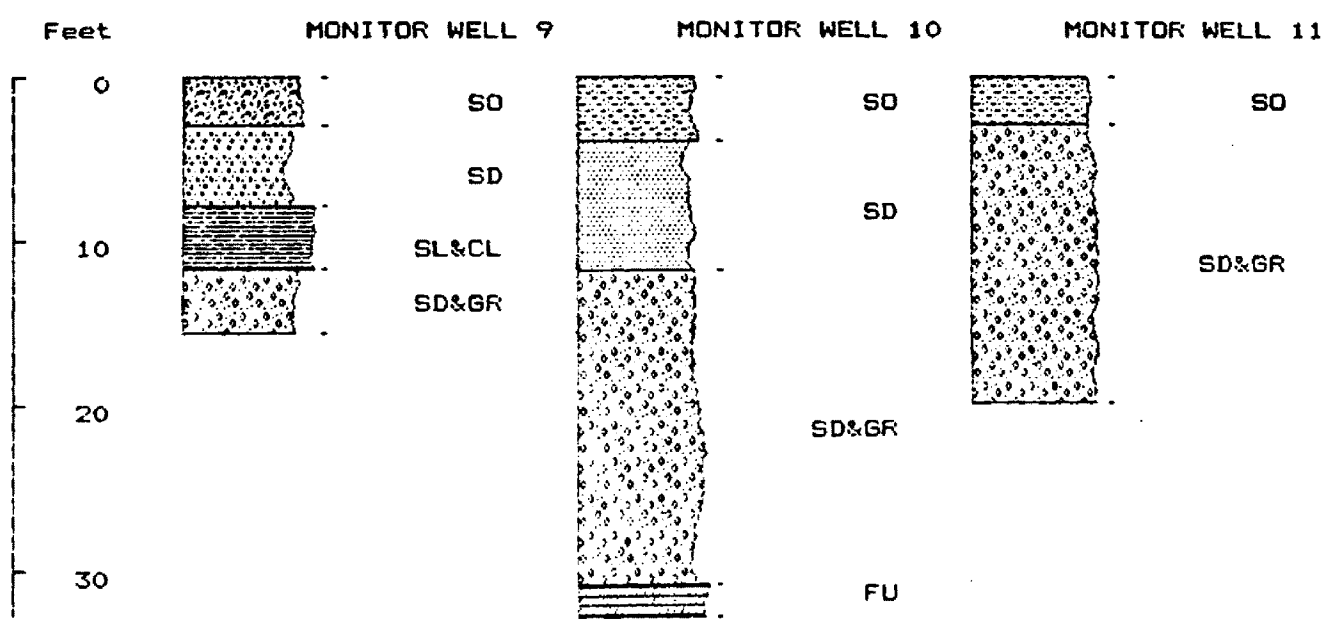
ETNAW T-110bites STRATCOL

Scale 1: 120



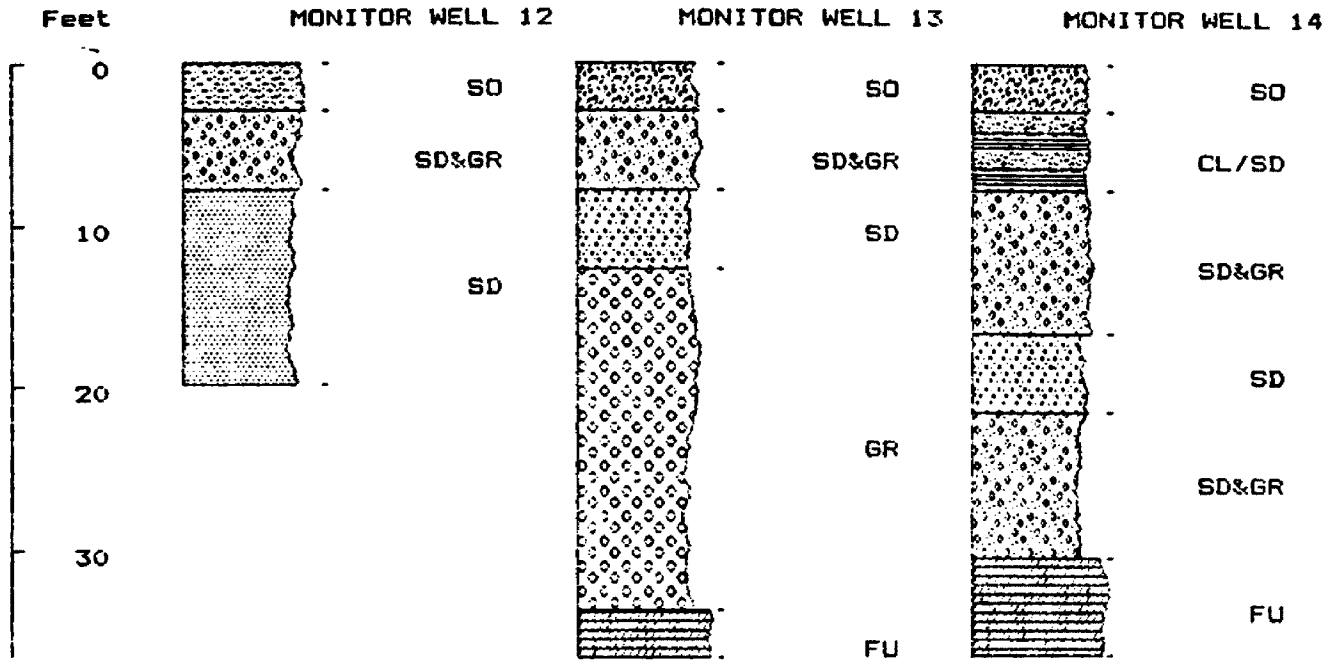
ETNAW T-110bites STRATCOL

Scale 1: 120



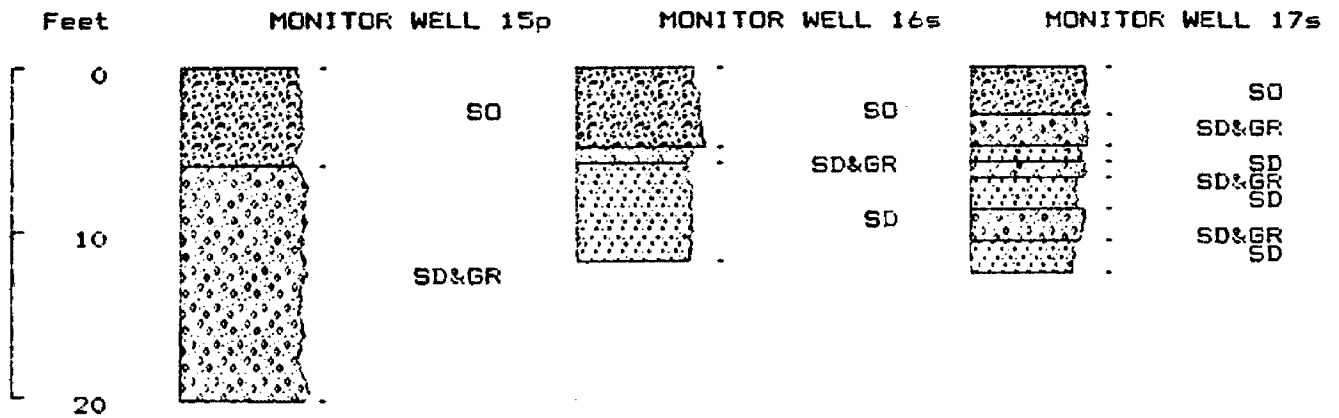
ETNA TRILLOBITES STRATCOL

Scale 1: 120



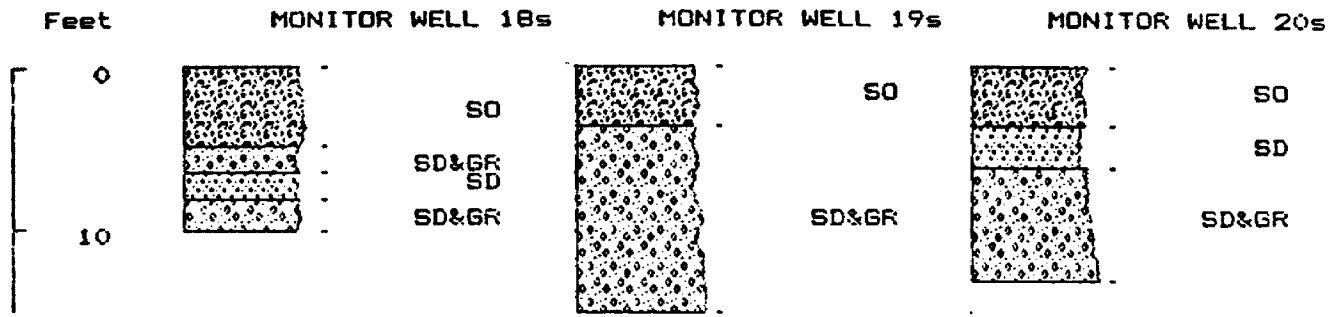
ETNA TRILLOBITES STRATCOL

Scale 1: 120



ETHAN T. LEBLANC STRATCOL

Scale 1: 120



SAND SAMPLE (MW 12)

SIEVE ANALYSIS OF SOIL AND AGGREGATE				FOREST —	
USDA FOREST SERVICE REGION 1				FIELD NO. SAND	LAB NO. —
				PROJECT IVERSON	
				LOCATION RICHLAND, CO.	
SIEVE SIZE	WEIGHT RETAINED	% RETAINED	% PASSING	SPEC. GRADATION	
3" (75.0mm)	0	0	100		
2" (50.0mm)	0	0	100		
1-1/2" (37.5mm)	0	0	100		
1" (25.0mm)	0	0	100		
3/4" (19.0 mm)	0	0	100		
1/2" (12.5 mm)	51.5 g	4.43	95.57		
3/8" (9.5 mm)	12.0 g	1.03	94.54		
#4 (4.75 mm)	44.0 g	3.78	90.76		
PAN	1055.0 g	90.75			
TOTAL	1162.0 g				
Sieve Loss or Gain	-3 g	WEIGHT OF ORIGINAL SAMPLE <u>1165.5 g</u>			
SIEVE SIZE	Weight Retained X $\frac{G}{H}$	= % RETAINED	% PASSING	SPEC. GRADATION	
#8 (2.36 mm)	13.6	3.15	87.61		
#10 (2.00 mm)					
#16 (1.18 mm)	13.8	3.20	84.41		
#30 (.600 mm)	20.1	4.66	79.75		
#40 (.425 mm)					
#50 (.300 mm)	39.9	9.25	70.5		
#100 (.150 mm)	139.0	32.24	38.26		
#200 (.075 mm)	109.4	25.37	12.89		
Total Minus #200	E 55.5	12.87	MINUS #40 MATERIAL IS: PLASTIC <input type="checkbox"/> NON-PLASTIC <input checked="" type="checkbox"/>		
TOTAL	H 391.3				
Sieve Loss or Gain	H-A .8				
TOTAL DRY WEIGHT BEFORE WASHING		A	390.5	REMARKS	
TOTAL DRY WEIGHT AFTER WASHING, BEFORE SIEVING		B	347.6		
MINUS #200 WASHED OUT (A-B)		C	42.9		
MINUS #200 FROM SIEVING PAN		D	12.6		
TOTAL MINUS #200 (C+D)		E	55.5		
MINUS #4 CORRECTION FACTOR, $\frac{G}{H}$.23194		
FORM NO. 51-7100 MTL-2	2/80	OPERATOR SP	DATE 11/87	CHECKED BY	DATE

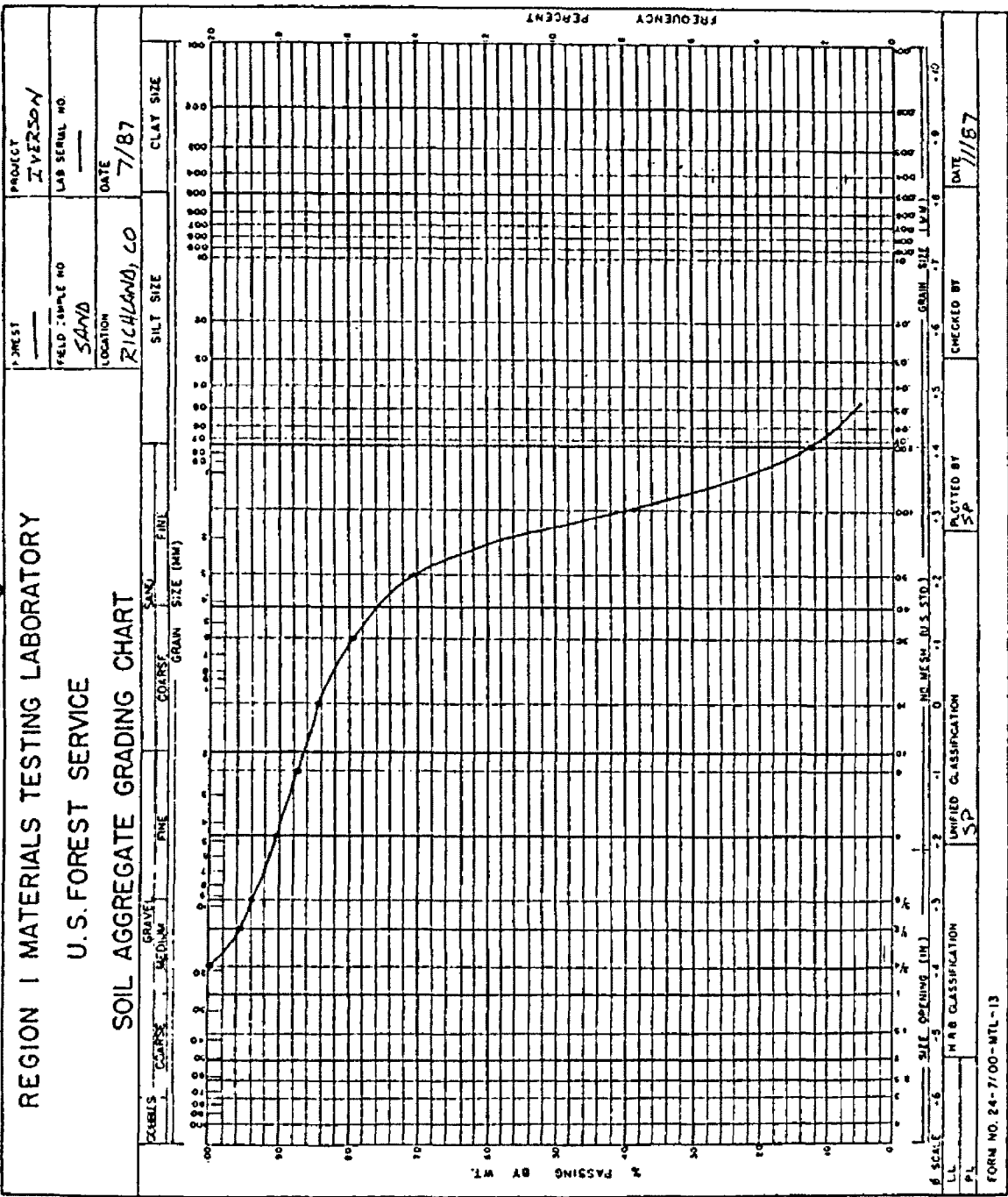
SAND AND GRAVEL SAMPLE

SIEVE ANALYSIS OF SOIL AND AGGREGATE		FOREST _____	
USDA FOREST SERVICE REGION 1		FIELD NO. SAND & GRAVEL _____	LAB NO. _____
		PROJECT <u>IVERSON</u>	
		LOCATION <u>RICHLAND, CO.</u>	
SIEVE SIZE	WEIGHT RETAINED.	% RETAINED	% PASSING
3" (75.0mm)	0	0	100
2" (50.0mm)	0	0	100
1-1/2" (37.5mm)	157.0 g	2.57	97.43
1" (25.0mm)	104.0 g	1.70	95.73
3/4" (19.0 mm)	264.0 g	4.31	91.42
1/2" (12.5 mm)	444.5 g	7.26	84.16
3/8" (9.5 mm)	287.5 g	4.70	79.46
#4 (4.75mm)	621.0 g	10.15	G 69.31
PAN	4242.0 g	69.31	
TOTAL	6120.0 g		
Sieve Loss or Gain	- 1 g		
		WEIGHT OF ORIGINAL SAMPLE <u>6121.0 g</u>	
SIEVE SIZE	Weight Retained X $\frac{G}{H}$	= % RETAINED	% PASSING
#8 (2.36 mm)	55.4 g	6.77	62.54
#10 (2.00mm)			
#16 (1.18 mm)	67.4g	8.23	54.31
#30 (.600mm)	95.1 g	11.61	42.7
#40 (.425 mm)			
#50 (.300mm)	84.1 g	10.27	32.43
#100 (.150 mm)	144.9 g	17.70	14.73
#200 (.075 mm)	59.9 g	7.31	7.42
Total Minus #200	E 60.7 g	7.41	
TOTAL	H 567.5 g		
Sieve Loss or Gain	H-A - .9 g		
			MINUS #40 MATERIAL IS: PLASTIC <input type="checkbox"/> NON-PLASTIC <input checked="" type="checkbox"/>
TOTAL DRY WEIGHT BEFORE WASHING A <u>568.4</u>			REMARKS
TOTAL DRY WEIGHT AFTER WASHING, BEFORE SIEVING B <u>513.8</u>			
MINUS #200 WASHED OUT (A-B) C <u>54.6</u>			
MINUS #200 FROM SIEVING PAN D <u>6.1</u>			
TOTAL MINUS #200 (C+D) E <u>60.7</u>			
MINUS #4 CORRECTION FACTOR, $\frac{G}{H}$ = <u>.12213</u>			
FORM NO. 51-7100 MTL-2	2/80	OPERATOR <u>SP</u>	DATE <u>11/87</u>
		CHECKED BY	DATE

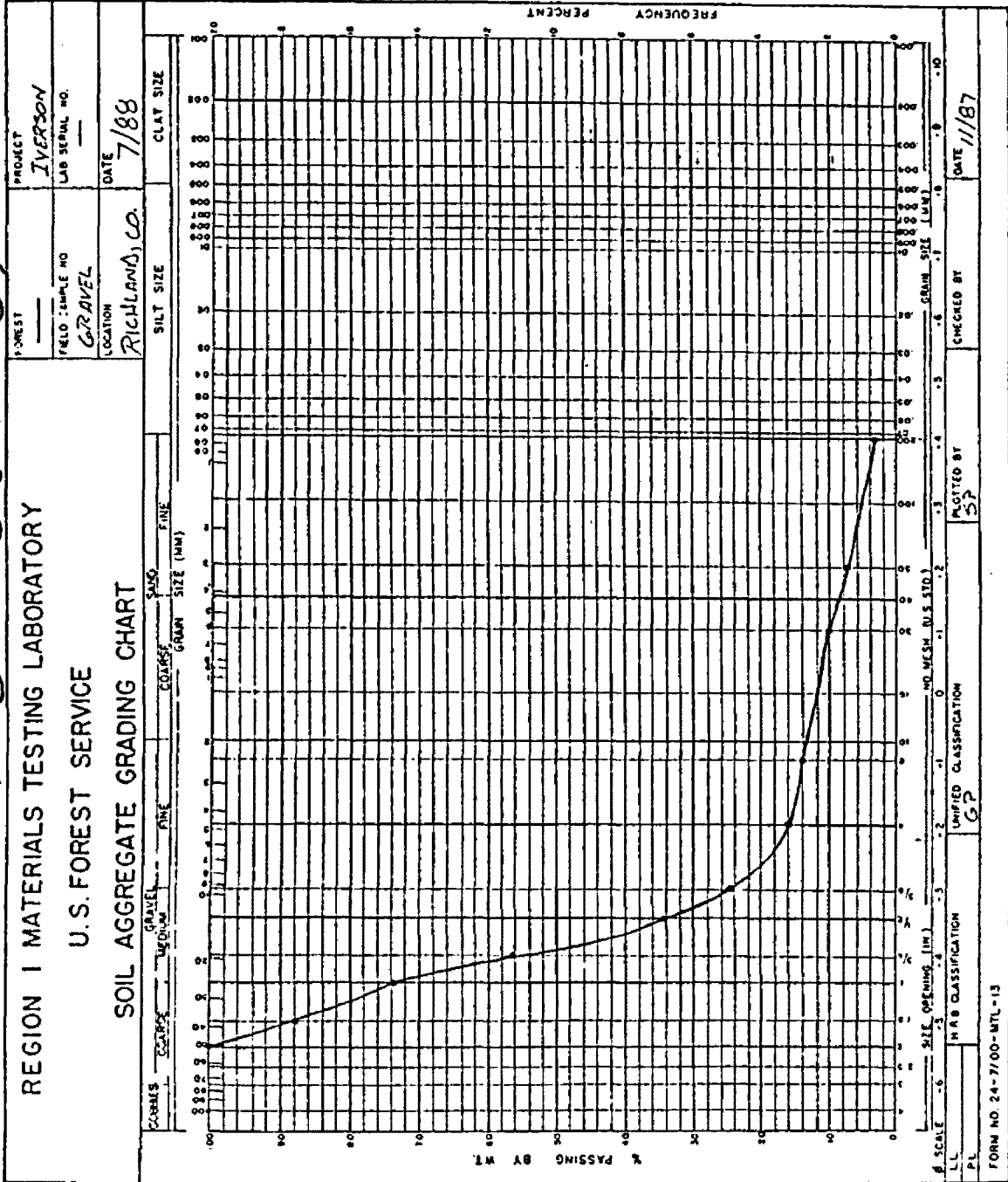
GRAVEL (MW 13) SAMPLE

SIEVE ANALYSIS OF SOIL AND AGGREGATE				FOREST —	
USDA FOREST SERVICE REGION 1				FIELD NO. GRAVEL	LAB NO. —
				PROJECT IYERSON	
				LOCATION RICHLAND, CO.	
SIEVE SIZE	WEIGHT RETAINED	% RETAINED	% PASSING	SPEC. GRADATION	
3" (75.0mm)	0	0	100		
2" (50.0mm)	0	0	100		
1-1/2" (37.5mm)	253.5 g	12.03	87.97		
1" (25.0mm)	305.5 g	14.49	73.48		
3/4" (19.0mm)	361.5 g	17.15	56.33		
1/2" (12.5mm)	464.0 g	22.01	34.32		
3/8" (9.5mm)	254.0 g	12.05	22.27		
#4 (4.75mm)	132.5 g	6.29	15.98		
PAN	337.5 g	16.01			
TOTAL	2108.0g				
Sieve Loss or Gain	- 11.5 g			WEIGHT OF ORIGINAL SAMPLE <u>2119.5 g</u>	
SIEVE SIZE	Weight Retained X $\frac{G}{H}$	= % RETAINED	% PASSING	SPEC. GRADATION	
#8 (2.36mm)	43.5 g	2.05	13.93		
#10 (2.00mm)					
#16 (1.18mm)	41.7 g	1.97	11.96		
#30 (.600mm)	51.1 g	2.41	9.55		
#40 (.425mm)					
#50 (.300mm)	39.5 g	1.86	7.68		
#100 (.150mm)	52.6 g	2.48	5.21		
#200 (.075mm)	42.0 g	1.98	3.23		
Total Minus #200	E 68.5 g	3.23		MINUS #40 MATERIAL IS: PLASTIC <input type="checkbox"/> NON-PLASTIC <input checked="" type="checkbox"/>	
TOTAL	H 338.9 g				
Sieve Loss or Gain	H-A - 1.4 g			REMARKS _____ _____ _____ _____ _____	
TOTAL DRY WEIGHT BEFORE WASHING	A 337.5				
TOTAL DRY WEIGHT AFTER WASHING, BEFORE SIEVING	B 276.8				
MINUS #200 WASHED OUT (A-B)	C 60.7				
MINUS #200 FROM SIEVING PAN	D 7.8				
TOTAL MINUS #200 (C+D)	E 68.5				
MINUS #4 CORRECTION FACTOR, $\frac{G}{H}$	= .047153				
FORM NO. 51-7100 MTL-2	2/80	OPERATOR SP	DATE 11/87	CHECKED BY	DATE

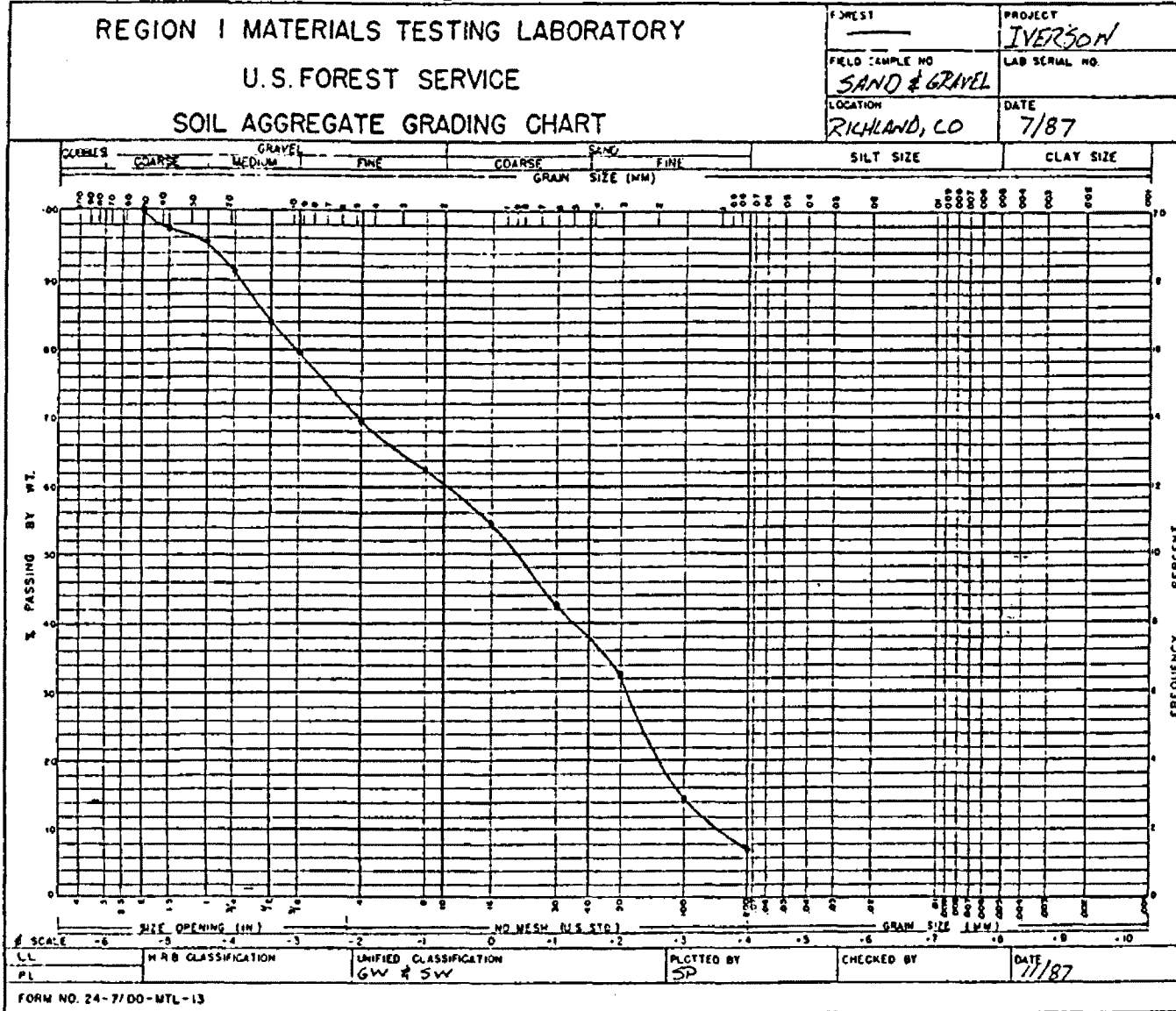
SAND (MW 12) SAMPLE

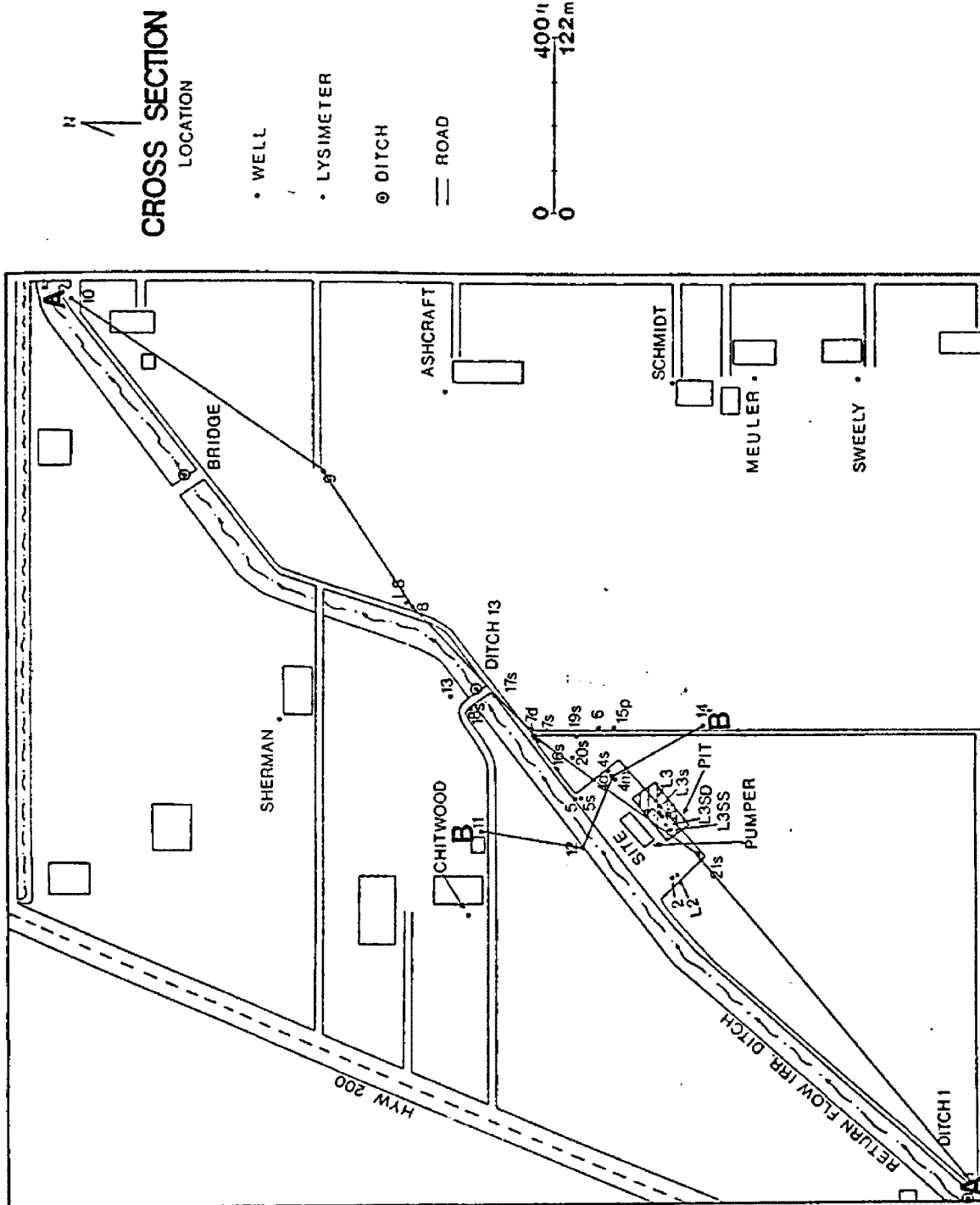


GRAVEL (M1W13) SAMPLE



SAND AND GRAVEL SAMPLE





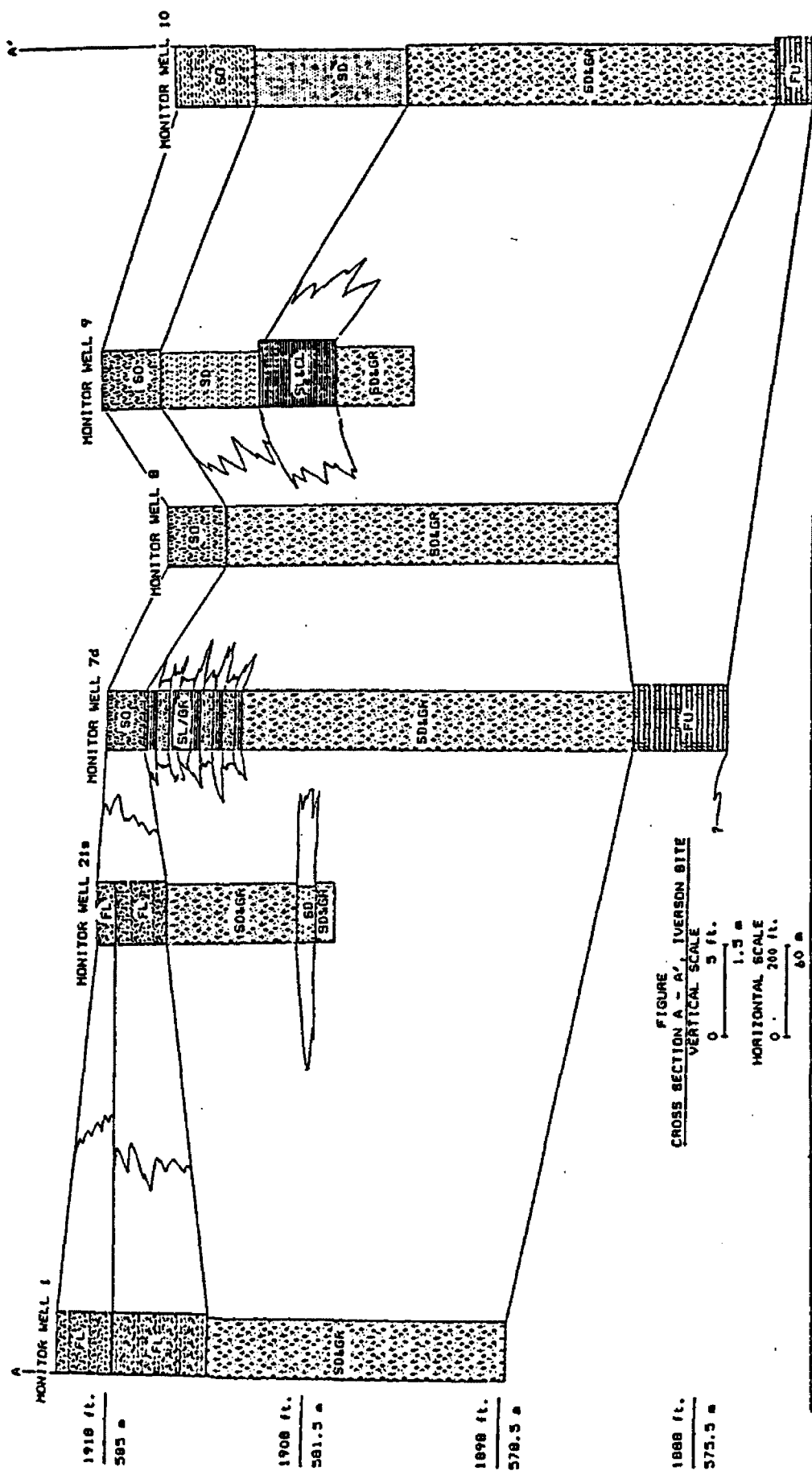
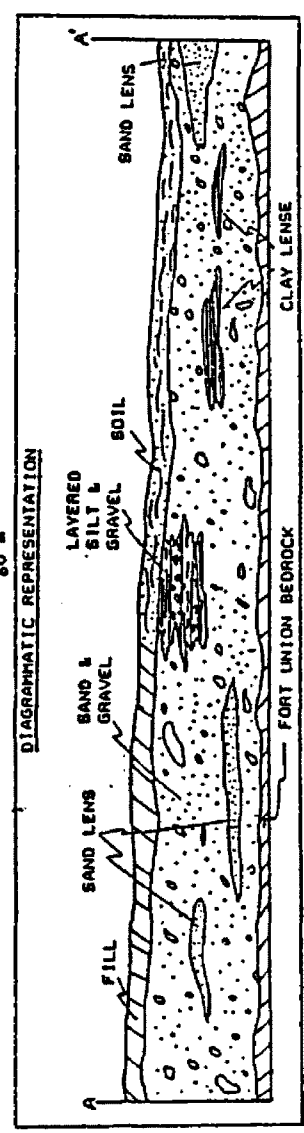
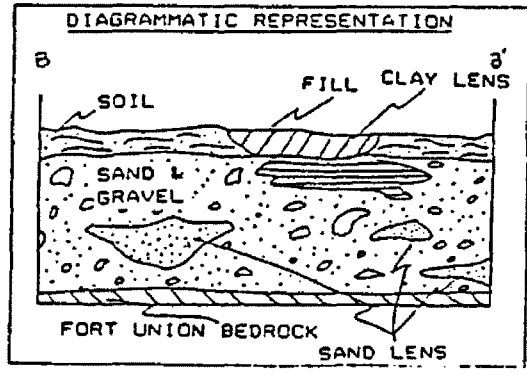
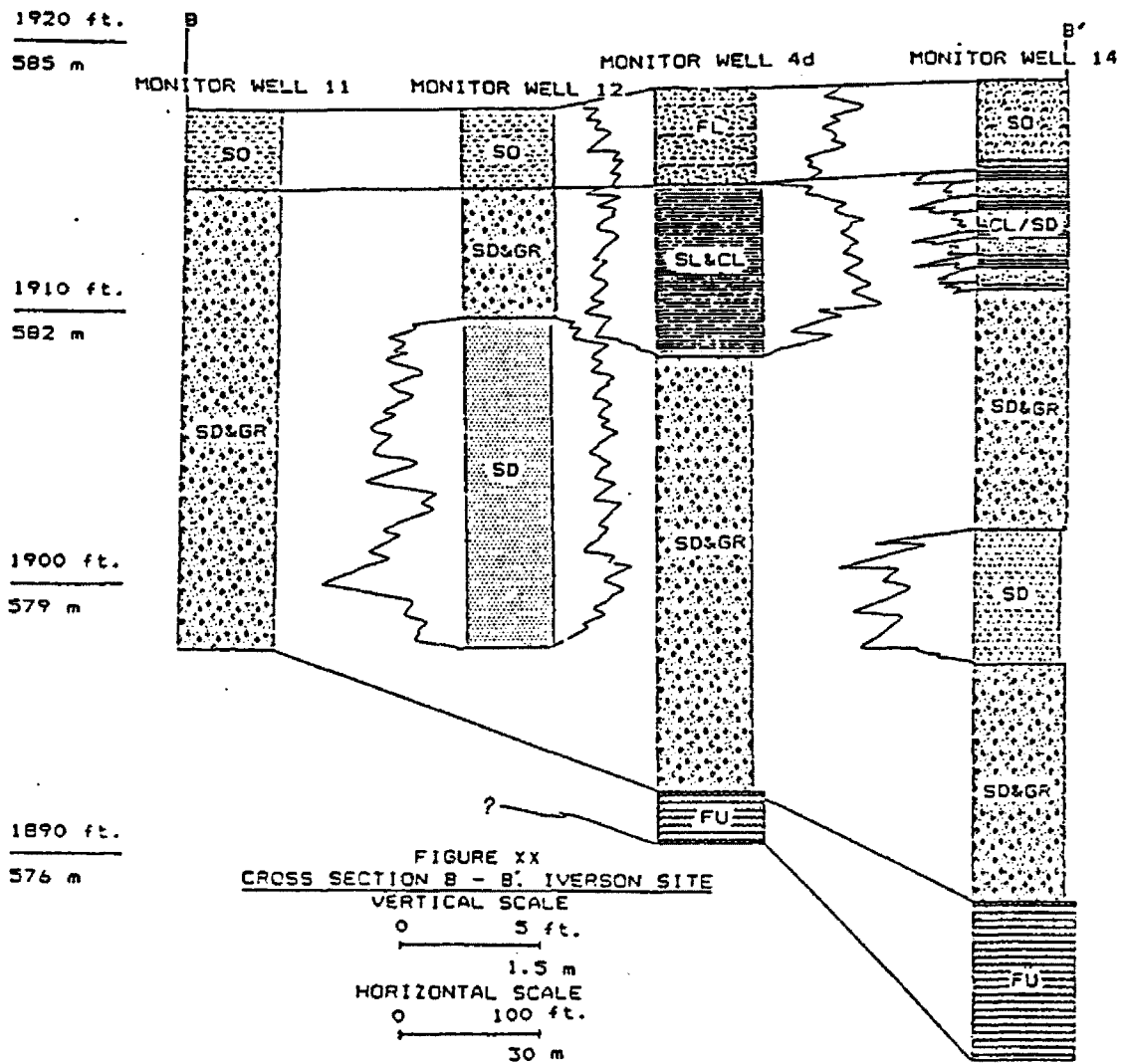


FIGURE
CROSS SECTION A - A', IVERSON SITE
VERTICAL SCALE
0 5 ft.
1.5 m
HORIZONTAL SCALE
0 200 ft.
60 m

CROSS SECTION
A - A'





CROSS SECTION B-B'

APPENDIX E
QUANTAB DISCUSSION

PERFORMANCE TIPS

Good Results may be expected using the QUANTAB according to the directions. Occasionally unusual results may occur. The following ideas will help resolve problems which may arise.

- Low Results:** Caused by incomplete extraction of salt, plugged wick, variable product. Resolved by more complete mixing, filtering, pin hole in signal, yellow area, and recalibration.
- High Results:** Caused by variable product or over easy extraction. Resolved by recalibration, or using 10% more water for extraction.
- Slow Reaction or No Completion:** Caused by sample plugging, sealed off signal, or compressed wick. Resolved by filtering and/or larger dilution, pin hole in signal, hot water.
- Inaccurate High Readings:** Readings above 7 on the QUANTAB scale will cover a broad, sometimes unacceptable spread. Always read results below 7.0 by choosing appropriate QUANTAB or diluting.
- Variable Results:** Non-representative samples or variable product. Be sure sample and mixture are correct. Recalibrate batch.

If technical questions arise, write or call Venture Systems, Ames Division, Miles Laboratories, Inc. 1127 Myrtle Street, P.O. Box 70 Elkhart, Indiana 46515 U.S.A. Telephone: 219/264-8534.

CALIBRATION TABLE

Be sure control number for table corresponds to bottle label, control number. This table is made from representative samples of the batch. Small variations from table may be noted within any batch. For greater accuracy recalibration by user is suggested.

QUANTAB CHLORIDE TITRATOR 1176						CONTROL NO. 0192119A		
Calibration at time of manufacture								
QUANTAB Reading	NaCl in solution as % NaCl ppm Cl ⁻ mg Cl ⁻ /L		QUANTAB Reading	NaCl in solution as % NaCl ppm Cl ⁻ mg Cl ⁻ /L		QUANTAB Reading	NaCl in solution as % NaCl ppm Cl ⁻ mg Cl ⁻ /L	
1.0	0.025	143	4.0	0.189	1130	7.0	0.61	3630
1.2	0.032	190	4.2	0.21	1230	7.2	0.65	3910
1.4	0.039	233	4.4	0.22	1340	7.4	0.70	4220
1.6	0.046	279	4.6	0.24	1450			
1.8	0.055	329	4.8	0.26	1570			
2.0	0.064	384	5.0	0.28	1700			
2.2	0.074	441	5.2	0.31	1840			
2.4	0.084	501	5.4	0.33	1980			
2.6	0.094	564	5.6	0.36	2140			
2.8	0.105	630	5.8	0.38	2310			
3.0	0.117	700	6.0	0.41	2490			
3.2	0.130	780	6.2	0.45	2680			
3.4	0.144	860	6.4	0.48	2900			
3.6	0.158	950	6.6	0.52	3130			
3.8	0.173	1040	6.8	0.56	3370			

11/15/79



Ames Division, Miles Laboratories, Inc. P.O. Box 70, Elkhart, Indiana 46515

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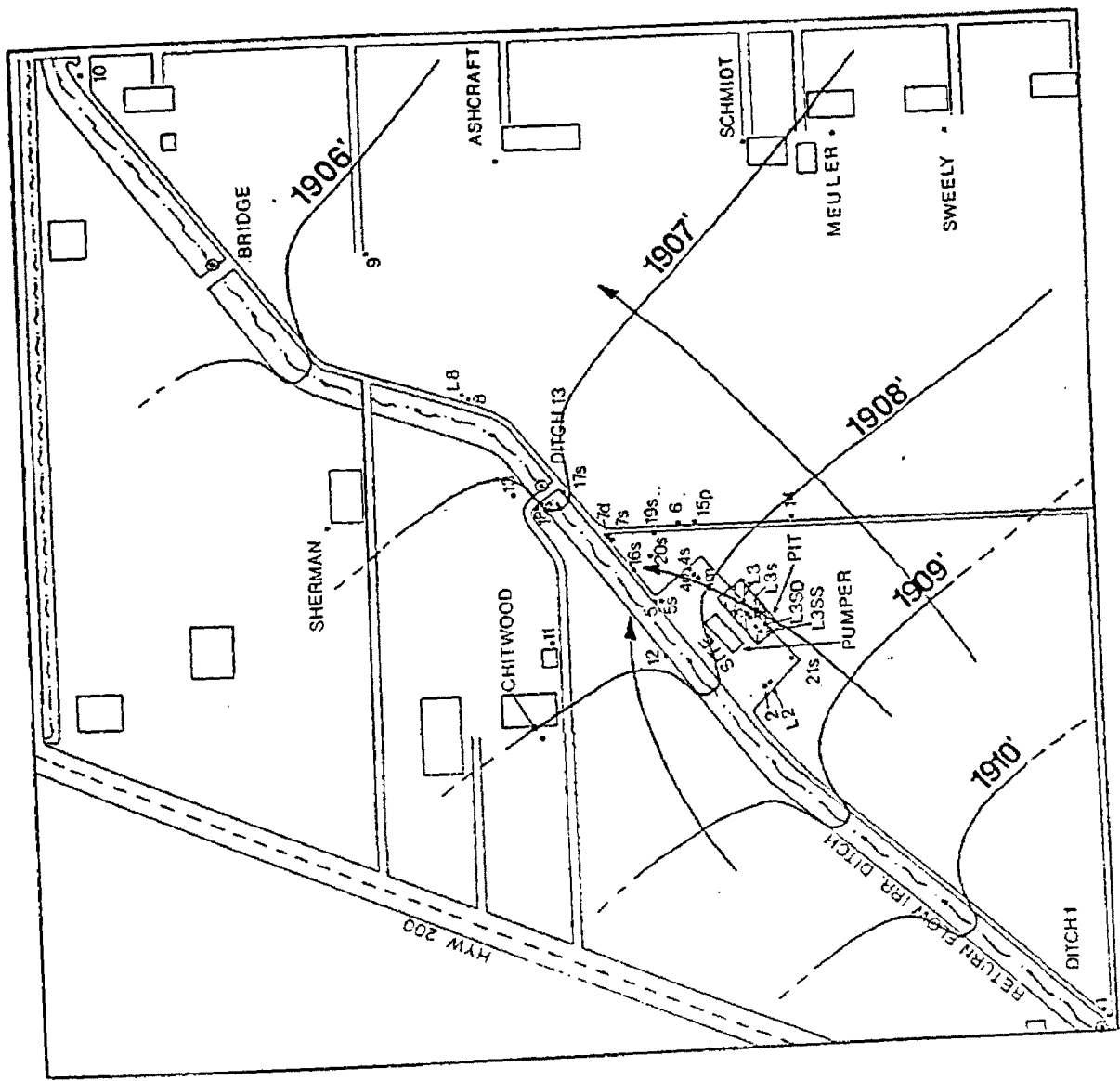
APPENDIX F
FALL AND WINTER GROUND WATER FLOW MAP

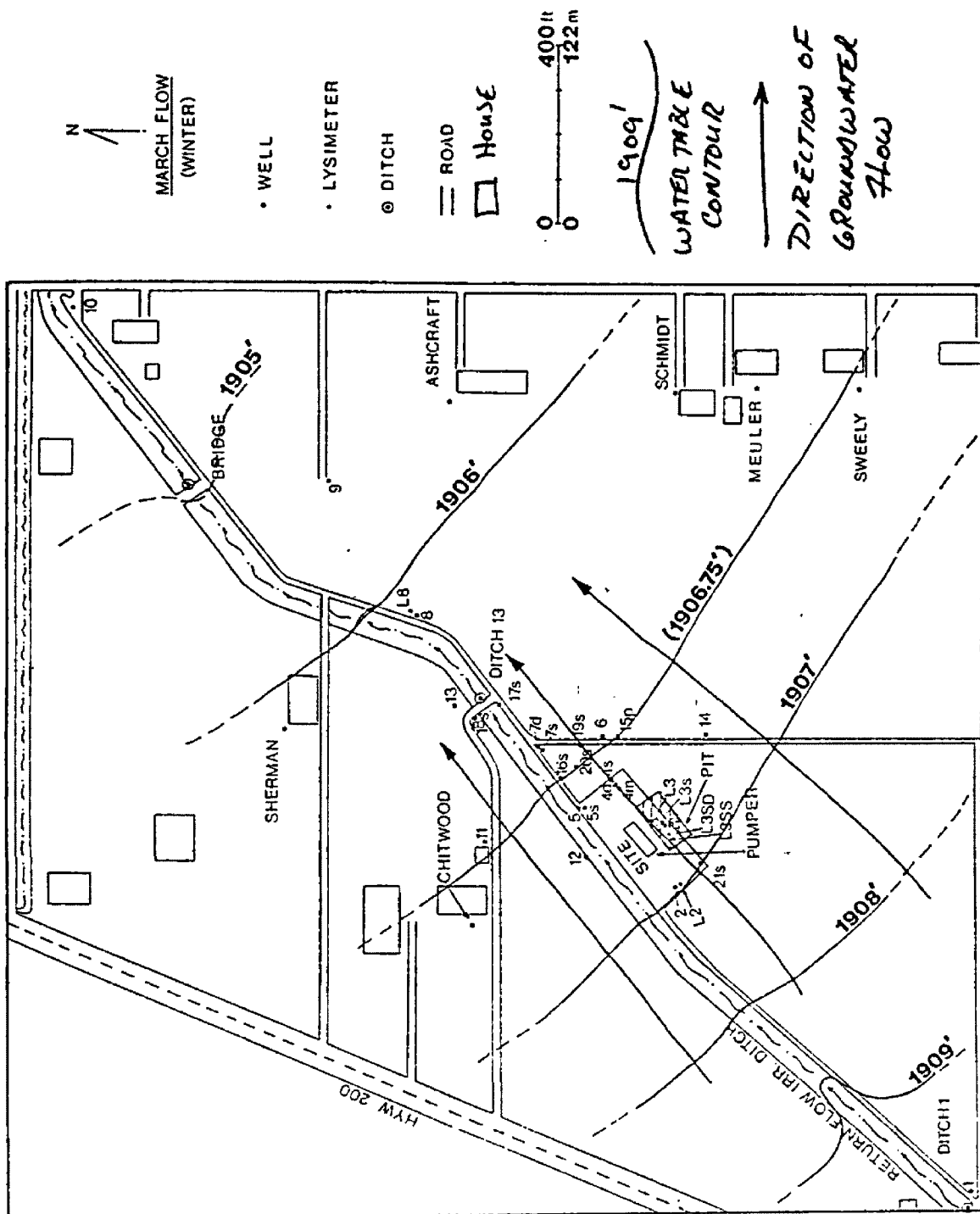
**OCTOBER FLOW
(FALL)**

- WELL
- LYSIMETER
- ⊙ DITCH
- == ROAD
- HOUSE



1910'
 WATER TABLE CONTOUR
 → DIRECTION OF
 GROUND WATER FLOW





APPENDIX G
CLIMATE DATA

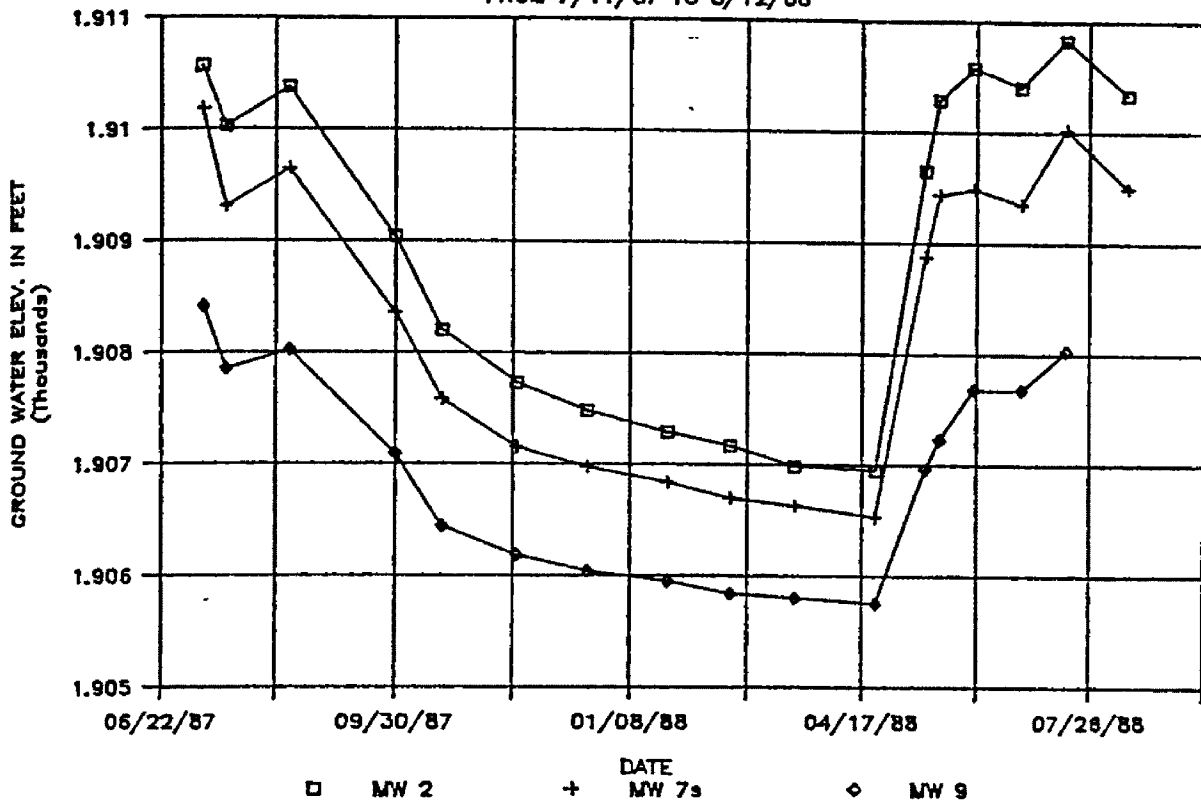
CLIMATE DATA FOR THE SIDNEY AREA (MONTANA STATE UNIVERSITY EXPERIMENT STATION)

DATE	AV. MAX T		AV. MIN T		AVERG. T		INCH PPT (cm)		INCH EVP (cm)	
	(F)	(C)	(F)	(C)	(F)	(C)				
Jun-87	84.6	29.22221	53	11.66666	68.9	20.49999	4.02	10.21075	7.71	19.58332
Jul-87	85.9	29.94444	56.7	13.72222	71.3	21.83333	2.87	7.289770	6.14	15.59553
Aug-87	80.5	26.94444	51.4	10.77777	66	18.99999	0.97	2.463790	5.11	12.97934
Sep-87	77	24.99999	43.3	6.27777	60.2	15.66666	0.92	2.336790	3.59	9.118563
Oct-87	60.3	15.72222	31	-0.55555	45.7	7.611110	0.1	0.253998	2.63	6.680173
Nov-87	50.2	10.11111	24.2	-4.33333	37.2	2.888888	0.21	0.533397		
Dec-87	37.3	2.944444	15.1	-9.38888	26.2	-3.22222	0.05	0.126999		
Jan-88	25.1	-3.83333	-1.1	-18.3888	12	-11.1111	0.56	1.422394		
Feb-88	32.9	0.499999	5.8	-14.5555	19.3	-7.05555	0.49	1.244595		
Mar-88	51.4	10.77777	22.8	-5.11111	37.1	2.833333	0.58	1.473194		
Apr-88	67.6	19.77777	28.6	-1.88888	48.1	8.944443	0.17	0.431798	4.97	12.6238
May-88	82.1	27.83333	46.4	7.999999	64.2	17.88888	0.79	2.006591	8.24	20.9296
Jun-88	94.7	34.83332	61.4	16.33333	78.1	25.61110	1.27	3.225787	10.44	26.5176

APPENDIX H
ANNUAL WATER TABLE FLUCTUATION

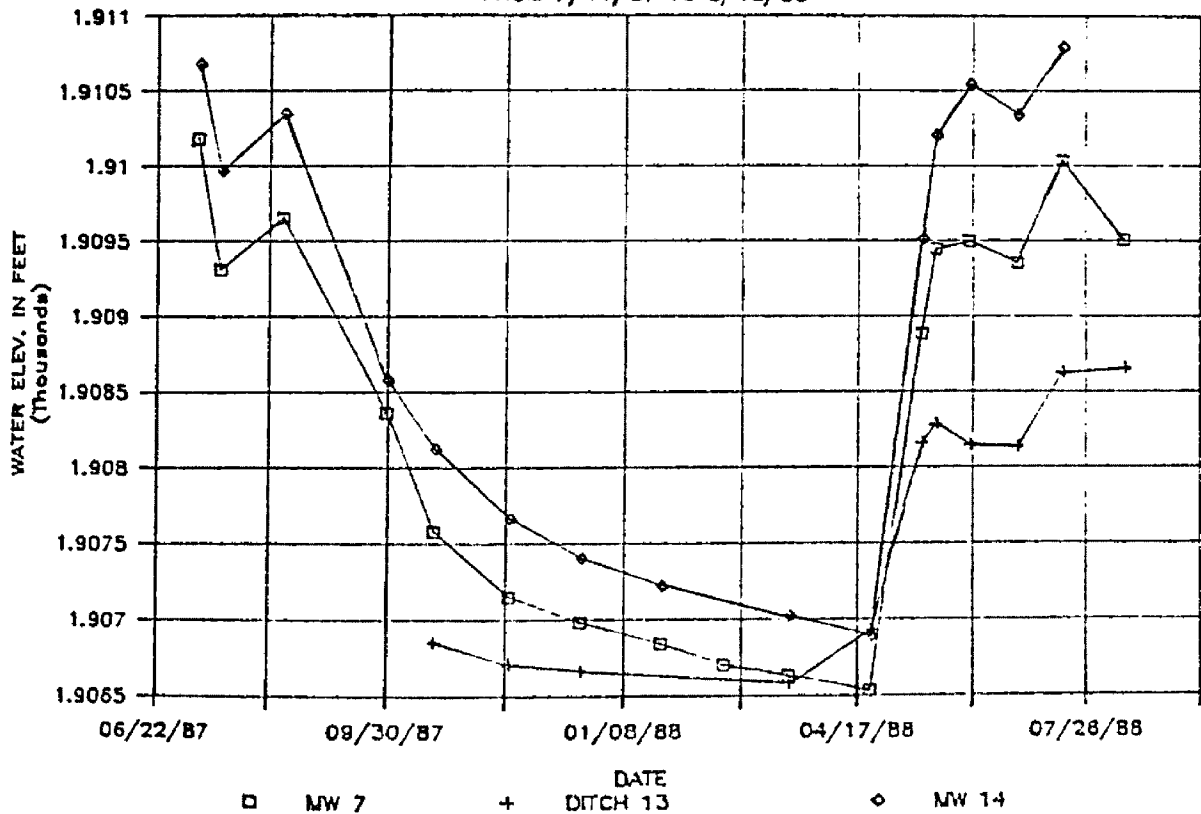
COMPARISON OF ANNUAL WL IN THREE WELLS

FROM 7/11/87 TO 8/12/88



COMPARISON OF DITCH WL AND WELL WL

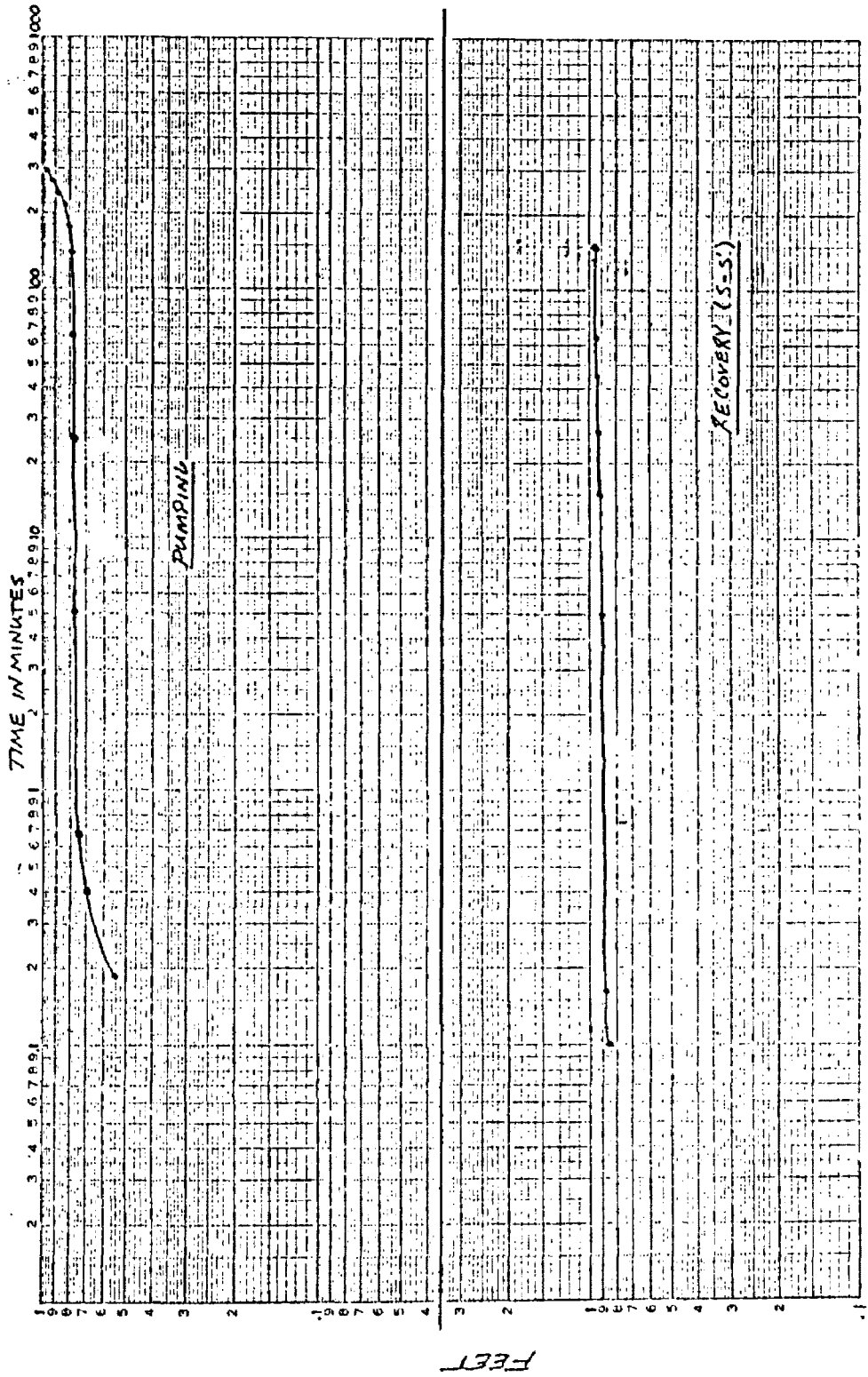
FROM 7/11/87 TO 8/12/88



APPENDIX I
AQUIFER TEST DRAWDOWN AND RECOVERY

PUMPING WELL 15B
 STATIC = 13.08'

Q = 30 GPM (SEE APPENDIX C)

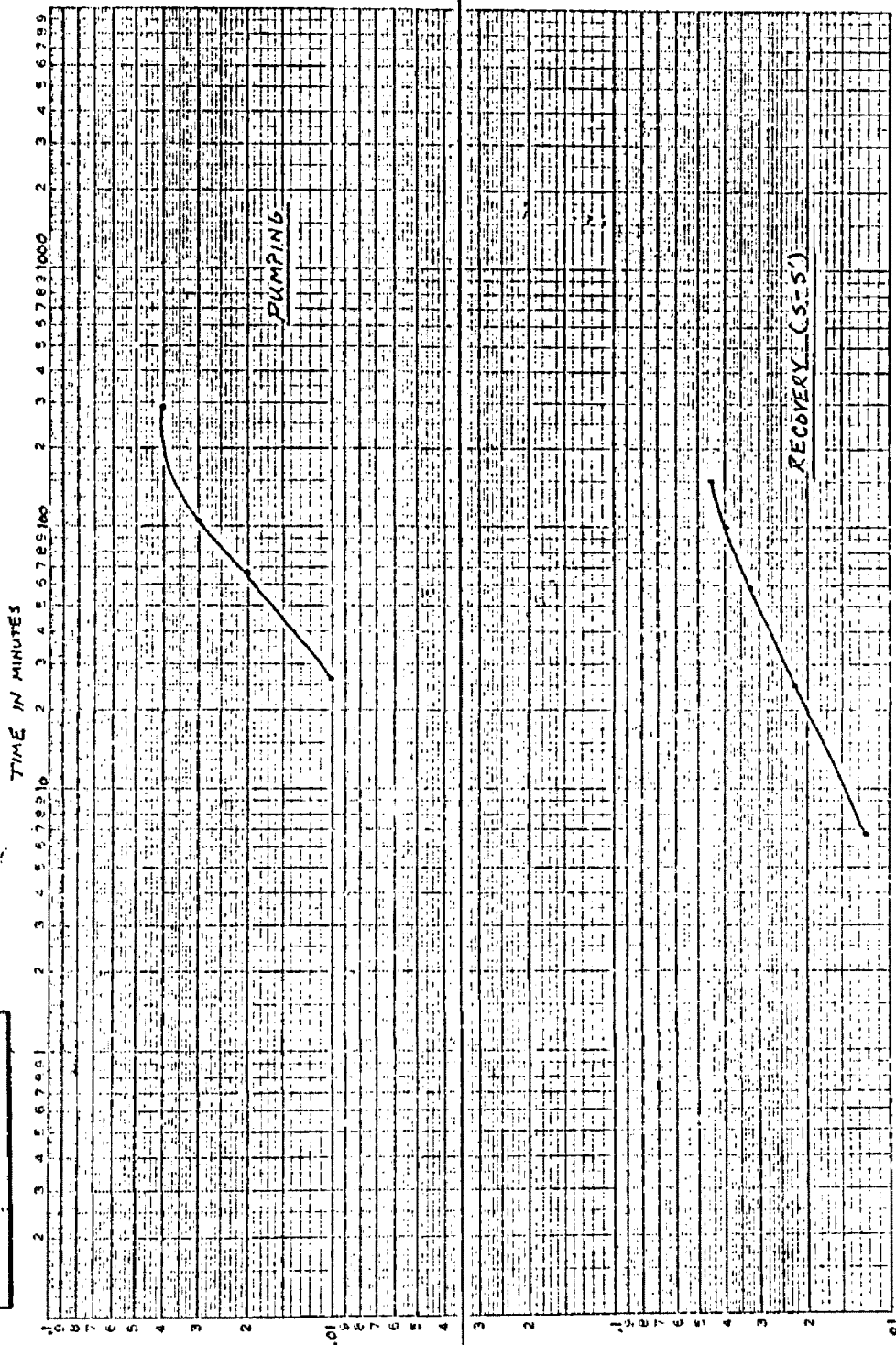


Full Logarithmic, 3 x 5

"Page 122 omitted in numbering."

MONITOR WELL 19S
 STATIC = 13.16'

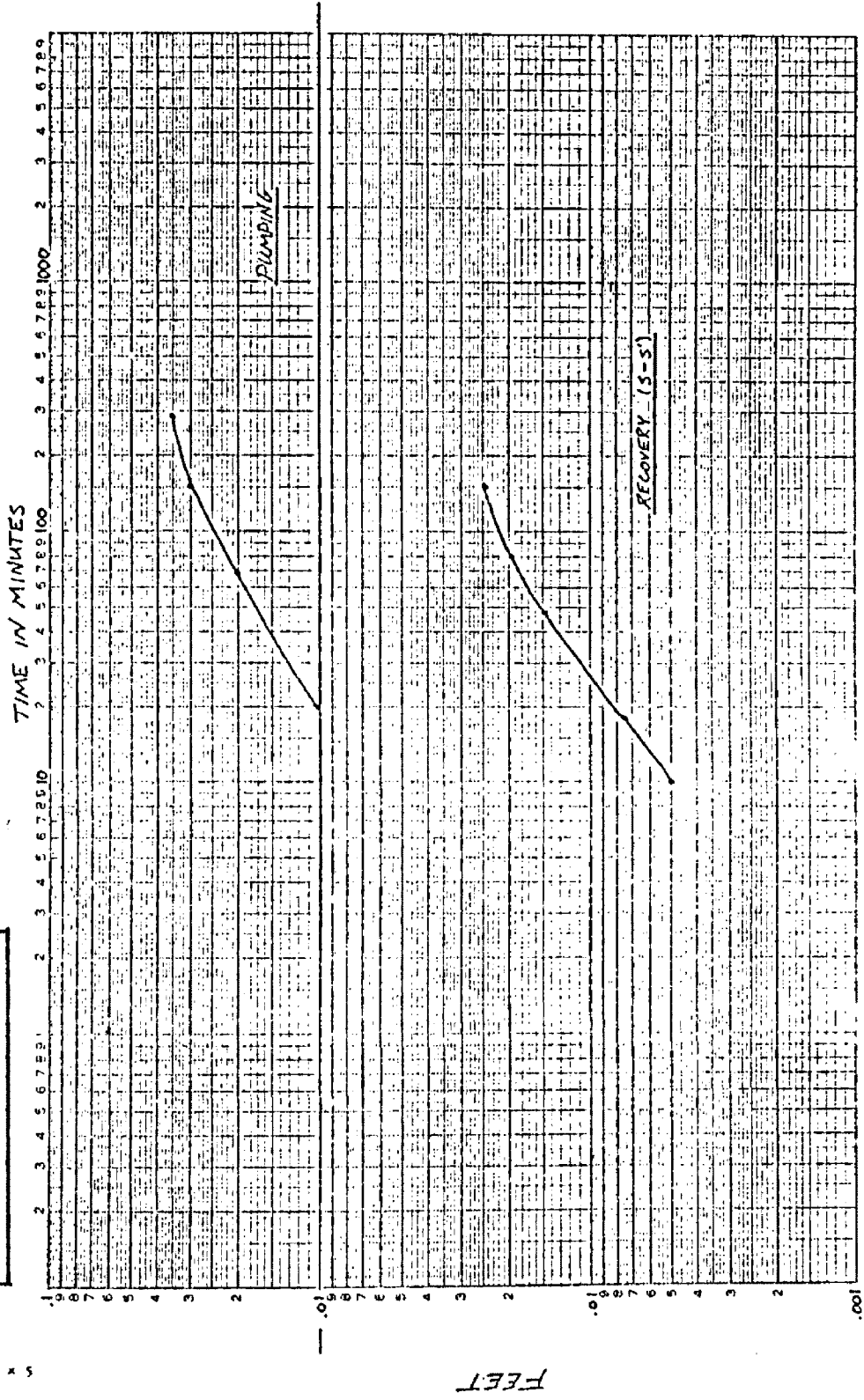
r = 82 ft.



Full Logarithmic, 3 x 5

MONITOR WELL ZOS
 STATIC = 15.565'

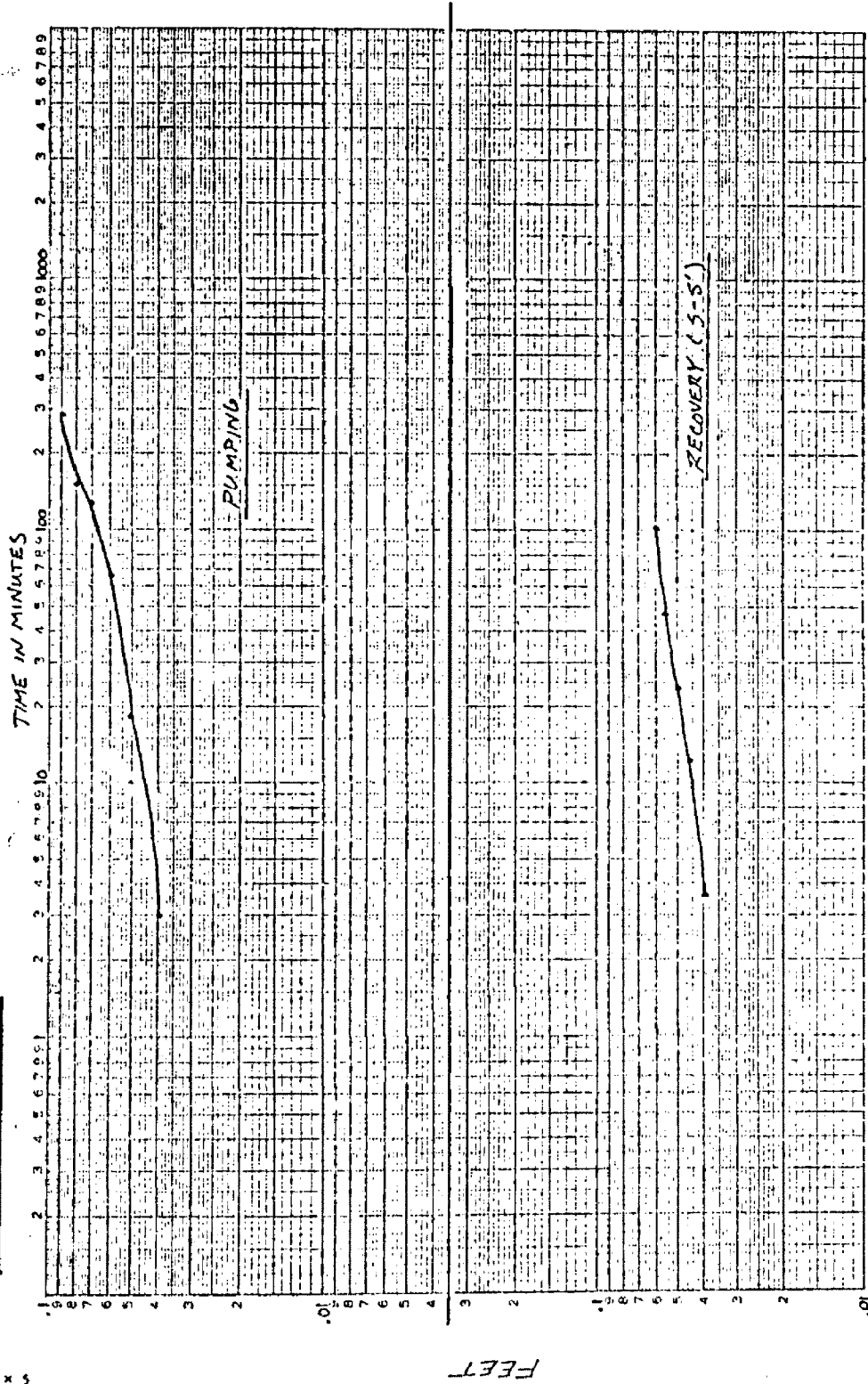
r = 118 ft.



Full Logarithmic, 3 x 5

MONITOR WELL 6
 STATIC = 13.52'

$r = 356t$



Full Logarithmic, 3 x 5

APPENDIX J
FIELD DATA

QUANTAS RESULTS OF ops C1 IN WATER AT THE IYERSON SITE, RICHLAND, CO.

Well #	DATE	07/11/87	07/21/87	08/17/87	10/20/87	11/22/87	12/19/87	01/23/88	02/21/88	03/20/88	04/23/88	05/24/88	05/26/88	09/12/88
1		<36	<36								50		<50	
2		<36	<36			<34							<50	30
21								41			34	<34	<50	
3		2447	2784	365					34	60	50		74	
3s				61	<34							1164	547	1212
4s		1230	1279	750								658	825	740
4d		126	36		34						50		<50	
4n		650	835	88	61	47	41	41	49	50			<50	
5		<36	52		34				41		50		<50	
5s				174	170						293	34	101	
6		<36	<36		<34						34		<50	
16				703	799			703	547	415	340	456	426	
20				750	547	425	425	365	277	241				
19				88	81	41		54	54	81	47	60		
15				<36	<34								<50	
7s		152	143	192	170	152	61	101	97	101	61	60		
7d		58	49		81						61		<50	
17				110	142	54			111	111	142	<34	<50	
18				115	122						87	<34	<50	
8		<36	36						47	55	60	<34	<50	
9		<36	<36								<34		<50	
10		<36	<36										<50	
11		<36	<36			34					50			
12		<<36	<36								34		<50	
13		53	59			111		54	54				<50	
14		<<36	<36								<34		<50	
ASHCRAFT		<<36	<36											
ECHNIGT														
MEULER		93	79											
SWEELY		<<36	<36								<34		<50	
LOMAN		<36												
SHERMAN														
CHITWOOD		<<36	<36											
L2		>6000	>6000	>6000										
L3s		>6000	>6000											
L3		>6000	>6000								>6000			
L8		<36	<<36											
DITCH 13		<<36	<36			<34	<34			<34			<50	30
DITCH 1	N/A		<36	<36	<34	<34	41	<34		<34	<34		<50	

CONDUCTANCE OF WATER IN ushos/cm AT THE INVERSON SITE, RICHLAND, CO. (CORRECTED TO 25°C)

Well #	DATE	07/11/87	07/21/87	08/17/87	10/20/87	11/22/87	12/19/87	01/23/88	02/21/88	03/26/88	04/23/88	05/24/88	05/26/88	06/12/88
1		1490	1597	1597	1563						1085		743	
2		925	1146	1061	1137	1096							999	823
2i					1190	1089		939			1026	1008	1057	
3		5250	6215	9914	2213				944	1057	1053	(4181)	1135	942
3e					1381	1173						4181	2351	3159
4e		6000	4635	4770	3245							2741	3206	2759
4d		1450	1631	1349	1579	1286					1170		1250	
4a		3000	3097	2768	1519	1289	1221	1209	1082	1065	1056	(2741)	1215	
5		900	1272	1309	1355	1242				1111			1187	937
5s					1572	1570							1487	947
5		1075	1461	1199	1192	1170							1092	990
16					3315	3317		2324	2339	1983	1729	2401	2284	1878
20					3528	2311	1999	1844	1910	1655	1655			
19					1417	1296	1117		1149	1170	1193	1195	1129	942
15					1170	1061							671	
7s		1500	1679	1730	1691	1683	1401	1250	1294	1237	1286	1121	1134	945
7d			1288	1317	961	1372					1028		1223	
17					1467	1479	1346		1307	1298	1302	746	807	811
18					1623	1593					1028	1254	927	1099
8		1200	1127	1220	1285	1157			1114	1146	1207	1031	456	
9		1100	1051	1127	1209	1115					1069		711	
10		950	893	1021	1105	1041							1055	
11		1300	1311	1541	1644	1485					1383			
12		810	672	1146	1226	1257					1190		1012	
13		1200	1290	1438	1392	1481			1017	1174			1050	
14		900	710	967	1125	1086					1046		1000	
ASHCRAFT			834	914										
SCHMIDT				723	536	467								
MEULER			1525	1534										
SHEELY			814	962	1044	990					937		850	
LOMAN			918											
SHERMAN														
CHITWOOD			1302	1372	1523	1542								
L2			15894	29897										
L3e			250000											
L3			250000	250000		250000					250000			
L8			3720	3100										
DITCH 13			946	904	1468	1339	1182			1085			876	735
DITCH 1				677	1521	1413	1681	1142		1012	861		759	

TEMPERATURE OF WATER IN CENTIGRADE AT THE IVERSON SITE, RICHLAND, CO.

Well #	DATE	07/11/87	07/21/87	08/17/87	09/25/87	10/20/87	11/22/87	12/19/87	01/23/88	02/21/88	03/20/88	04/23/88	05/24/88	06/26/88	08/12/88
1		14.4	12.9	14.8		12.5						7.1		16	
2		10.0	10.9	12.2		9.5	6.5							13.9	16.9
21						10.5	8.9		9.9			8	11.9	13.9	
3		15.6	13.0	14.2		12				5.9	6.9	7.9		15.9	16.9
3s						12	9						11.9	15.1	16.1
4s		12.8	13.2	15.0		12							13.3	13.5	15.9
4d		12.8	10.8	13.8		10.5	9.1					7.9		11.9	
4e		17.8	11.8	12.8		10.5	9	11.1	11.5	7.1	7	8		13.5	
5		11.1	12.9	13.8		10	9.1			6.1			8.1	11.1	
5s						10.5	9.2						8	11.1	12.3
6		11.7	12.4	12.5		11	9.1					7.1		13.2	
16						10.5	9		11	5	8.9	8.9	11.1	13.1	15.1
20						10	9.1	10.1	10.1	5.9	6.9	7.9			
19						11	9.1	10.9		6.9	7.5	7.9	10.1	15.2	16.9
15						10.5	9.9							15.9	
7s		11.7	12.9	12.0		11	9.9	9.9	8.1	6.9	7	9.1	10.9	12.9	17
7d			12.4	11.1		10.5	9.1					6.9		11.9	
17						11.5	8.2	8.2		5	7	6.9	15.1	17.1	18.1
18						12	9					6.5	13.1	15.9	
8		14.4	13.9	12.9		11	10			6	7	7	11	18.9	
9		22.2	13.1	13.9		11.5	10					7		17.9	
10		13.9	13.9	13.0		11.8	8							15.1	
11		12.2	12.1	12.9		11.1	9.9					6.9			
12		11.1	12.7	13.9		10.5	9.1					8		14.9	
13		23.9	14.0	16.1		11.5	8.9			8.9	6.1			15.9	
14		12.2	16.5	15.0		13	11					6.9		14.9	
ASHCRAFT			11.9	12.5											
SCHMIDT				15.0		13	11.1								
MEULER			14.2	14.3											
SWEELY			14.8	16.0		13.6	9.9					6.8		15.1	
LOMAN			15.8												
SHERMAN															
CHITWOOD			24.9	19.1		11.1	9.9								
L2			17.0	20.0											
L3s			18.2	20.5											
L3			16.7	17.0			5.2								
LB			18.0												
DITCH 13			16.7			8.5	6	3.9				7		15.1	21.1
DITCH 1				15.1		9	6	3.9	5.1			4	12.5	23.1	

PH OF WATER AT THE IVERSON SITE, RICHLAND, CO.

Well #	DATE	07/11/87	07/21/87	08/17/87	09/25/87	10/20/87	11/22/87	12/19/87	01/23/88	02/21/88	03/20/88	04/23/88	05/24/88	06/26/88	08/12/88
1			7.7			6.9									7.1
2			5.3			7	6.89								
21						7	6.7		7.1						7.5
3			6.8			7.2				6.4	6.9				7.1
3s						6.8	6.1								
4s			4.5			6.8									
4d			6.7			7.3	7								7.3
4m			4.8			7.3	6.4	7.49	7.1	6.8	7.2				7.3
5			6			7.1	7.2			7.2					7.3
5s						7	6.8								7.2
6			6.3			7.1	7.1								7.1
16						7.3	7		6.8	6.9	7.4				7.4
20						7.3	7	7.41		6.7	6.6				7.1
19						7.1	6.9	6.6		6.8	6.9				7.1
15						7.2	7.2								
7s			4.9			7.2	7.2	6.9	7.1	6.8	6.7				7.1
7d			5.5			7.4	7.5								7.9
17						7.1	7.1	6.9		6.8	7				7.2
18						7.2	7.3								7.4
8			5.3			7.2	6.7			7.1	7.1				7.3
9			7.5			7.15	6.9								7.3
10			7.3			7.2	6.8								
11			5.5			7.2	6.9								7.2
12			6.5			7.2	7								7
13			5.5			7.2	6.9			7.3	7.2				
14			5.8			7.1	6.9								7.3
ASHCRAFT			7.7												
SCHMIDT															
MEULER			7												
SWEELY		8	6.8			7.5									6.8
LDHAN		8													
SHERMAN															
CHITWOOD			7.8												
L2		6.9	6.4			6.7									
L3s		6.0	6.6												
L3		9.5	6.1												
L8		7.6	7.6												
DITCH 13			5.5			8.1	7.9	7.9			8.1				
DITCH 1	N/A		8			7.8	6.81	7.7	7.5		6.8				8

DEPTH BELOW GROUND SURFACE TO THE WATER TABLE AND GROUND WATER ELEVATION AT THE IVERSON SITE, RICHLAND, SC.

WELL	MP ELEV.	MEM	MP	DATE	ELEV.	MEM	MP	DATE	ELEV.	MEM	MP	DATE	ELEV.	MEM	MP	DATE	ELEV.	MEM	MP	DATE	ELEV.	
1	1922.1	8.82	8.60	07/11/87	10/20/87	11/21/87	12/21/87	01/24/88	02/20/88	03/19/88	04/23/88	05/15/88	05/21/88	06/05/88	06/26/88	07/16/88	08/12/88					
2	1920.14	9.57	10.11	09/13/87	10/10/87	11/10/87	12/10/87	01/10/88	02/10/88	03/10/88	04/10/88	05/10/88	06/10/88	07/10/88	08/10/88	09/10/88	10/10/88					
21	1920.64	1910.57	1910.03	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
3	1920.14	9.65	10.20	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
3s	1920.89	1910.49	1909.94	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
4s	1921.09	9.75	11.33	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
4d	1920.24	1910.34	1909.76	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
4a	1919.74	9.41	9.94	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
5	1919.59	8.99	9.92	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
5s	1919.34	1910.61	1909.67	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
6	1921.14	10.81	11.59	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
16	1919.87	1910.33	1909.55	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
20	1917.96	1907.72	1907.25	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
19	1920.74	13.06	13.47	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
15	1919.84	1907.69	1907.27	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
7s	1919.84	9.66	10.53	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
7d	1919.69	9.77	10.30	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
17	1918.18	1910.12	1909.39	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
18	1917.75	1907.32	1906.95	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					
8	1915.78 (1915.72)	1909.24	1908.79	09/09/87	10/09/87	11/09/87	12/09/87	01/09/88	02/09/88	03/09/88	04/09/88	05/09/88	06/09/88	07/09/88	08/09/88	09/09/88	10/09/88					

WELL LOCATION	MP ELEV.	NEW MP DATE	ELEV.	07/11/87	07/21/87	08/17/87	09/30/87	10/20/87	11/21/87	12/21/87	01/24/88	02/29/88	03/19/88	04/23/88	05/15/88	05/21/88	06/05/88	06/26/88	07/16/88	08/12/88	
9	1920.48		12.07	12.63	12.46	13.39	14.04	14.3	14.44	14.53	14.64	14.68	14.73	14.73	13.51	13.25	12.8	12.8	12.8	12.8	12.45
HL ELEV.	1908.41	1907.85	1908.02	1907.09	1906.44	1906.18	1906.04	1905.95	1905.84	1905.84	1905.75	1906.97	1907.23	1907.23	1906.97	1907.23	1907.68	1907.68	1907.68	1908.03	1908.03
10	1916.67		9.22	9.92	9.80	10.89	11.65	11.85	11.94	11.67	12.05	12.1	12.07	11.06	10.88	10.84	10.84	10.84	10.84	10.63	10.11
HL ELEV.	1907.45	1906.75	1905.87	1905.78	1905.02	1904.82	1904.73	1905	1904.51	1904.57	1904.6	1905.61	1905.79	1905.83	1906.04	1906.56	1906.56	1906.56	1906.56	1906.56	1906.56
11	1919.74		9.42	10.36	9.98	11.29	12.06	12.49	12.7	12.82	12.95	13.04	13.16	10.86	10.34	10.16	10.16	10.16	10.16	9.45	9.45
HL ELEV.	1910.32	1909.38	1909.36	1908.45	1907.68	1907.25	1907.04	1906.79	1906.7	1906.58	1906.88	1909.4	1909.58	1909.58	1910.29	1910.29	1910.29	1910.29	1910.29	1910.29	1910.29
12	1917.77		7.12	8.09	7.74	9.04	9.85	10.28	10.51	10.48	10.8	10.89	11	8.41	7.81	7.91	7.91	7.91	7.91	7.37	7.37
HL ELEV.	1910.65	1909.68	1910.03	1908.73	1907.92	1907.49	1907.26	1907.09	1906.97	1906.98	1906.77	1909.36	1909.96	1910.06	1909.86	1910.4	1910.4	1910.4	1910.4	1910.4	1910.4
13	1919.09		7.59	10.35	10.09	11.28	12.01	12.37	12.56	12.65	12.79	12.86	12.95	11	10.57	10.41	10.29	10.29	10.29	9.54	9.54
HL ELEV.	1911.5	1908.74	1909	1907.81	1907.08	1906.72	1906.53	1906.44	1906.3	1906.23	1906.14	1908.09	1908.52	1908.68	1908.8	1909.55	1909.55	1909.55	1909.55	1909.55	1909.55
14	1921.54		10.87	11.58	11.20	12.57	13.42	13.88	14.14	14.32	14.52	14.65	12.03	11.34	11	11.2	10.75	10.75	10.75	10.75	10.75
HL ELEV.	1910.67	1909.98	1910.34	1908.57	1908.12	1907.66	1907.4	1907.22	1907.02	1906.89	1909.51	1910.2	1910.54	1910.34	1910.79	1910.79	1910.79	1910.79	1910.79	1910.79	1910.79
ASHCRAFT	1919.37		11.5	12.97	13.28	13.43	13.55	13.64	13.66	13.75	13.84	13.86	13.75								
HL ELEV.	1907.87		1906.4	1906.09	1905.94	1905.82	1905.73	1905.71	1905.62												
SCHMIDT	1917.45		9.53	9.25	10.12	10.61	10.95	11.1	11.12	11.24	11.29	11.41	10.42	9.89							
HL ELEV.	1907.92	1908.2	1907.33	1906.82	1906.5	1906.35	1906.21	1906.16	1906.04	1907.03	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57	1907.57
NEULER	1918.47		9.83	10.73	11.22	11.57	11.72	11.74	11.83	11.96	12.11	12.11	12.11	10.24	9.5						9.52
HL ELEV.	1908.54	1907.74	1907.25	1906.9	1906.75	1906.73	1906.64	1906.51	1906.36												
SWEELY	1917.99		9.15	10	10.45	10.81	10.98	10.9	11.01	11.14	11.41	11.41	11.41	11.21	9.18	9.5	8.54	8.54	8.54	8.54	8.54
HL ELEV.	1908.64	1907.99	1907.54	1907.18	1907.01	1907.09	1906.98	1906.85	1906.59												
SHEPHERD	1918.49		9.61	9.58																	
HL ELEV.	1908.88	1908.91																			
CHITWOOD	1918.43		9.76	8.98																	
HL ELEV.	1908.67	1910.35																			
DITCH 13	1907.81																				
1908.36																					
1913.99																					
HL ELEV.	1906.85	1906.7	1906.66																		
BRIDGE	1913.51		6.75	7.75	8.31	8.45	8.46	8.63	8.65	8.6	8.16	7.96	7.79	7.95	7.84	6.86	6.86	6.86	6.86	6.86	6.86
HL ELEV.	1906.75	1905.76	1905.2	1905.06	1905.05	1904.88	1904.86	1904.91	1905.35	1905.55	1905.72	1905.56	1905.67	1906.65	1906.65	1906.65	1906.65	1906.65	1906.65	1906.65	1906.65
DITCH 1	1918.3 (1918.3)		5.56	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX	XX
1915.63			XX	3.55	3.92	3.98	4.24	4.15	3.85	3.62	3.5	2.95	3.45	3.72	3.12	3.12	3.12	3.12	3.12	3.12	3.12
HL ELEV.	1912.74	1912.08	1911.71	1911.65	1911.64	1911.48	1911.78	1912.01	1912.13	1912.68	1912.18	1911.91	1912.51	1912.51	1912.51	1912.51	1912.51	1912.51	1912.51	1912.51	1912.51

APPENDIX K
METAL ANALYSES
(UNIVERSITY OF MONTANA ICP)

DISSOLVED METALS

(NEGATIVE INDICATES LESS THAN DETECTION LIMIT).

METAL ANALYSES FOR WATER SAMPLES AT THE EVERSON SITE, RICHLAND, CO.

LOC.	DATE	ppm Al	ppm As	ppm Ca	ppm Cd	ppm Cu	ppm Fe	ppm K	ppm Hg	ppm Mn	ppm Na	ppm Ni	ppm Pb	ppm Sb	ppm P	ppm Si	ppm Ti	ppm Zn
HW 2	06/28/08	0.020	0.002	124.900	-0.001	0.001	0.010	3.600	29.700	0.001	64.500	0.001	-0.020	0.061	0.035	9.370	0.001	0.007
HW 3 (s) (SUP.)	10/22/07	0.000	0.003	104.800	0.001	0.000	0.010	6.500	35.500	0.001	233.300	0.005	-0.014	0.072	0.051	17.550	0.002	0.003
	04/23/08	0.020	0.006	97.600	-0.001	-0.001	0.010	3.600	39.000	0.300	110.000	0.006	-0.012	0.092	0.073	10.710	-0.001	0.001
	06/28/08	0.020	0.008	124.200	0.000	0.001	0.010	12.700	32.900	0.074	353.900	0.004	-0.007	0.089	0.042	12.360	0.003	0.001
HW 4 (s) (s)	10/22/07	0.110	0.017	112.900	0.000	0.002	0.030	11.600	35.100	0.409	218.200	0.008	0.010	0.080	0.130	12.710	0.002	0.005
	10/22/07	0.040	0.008	163.500	0.000	0.011	0.020	10.200	43.900	0.090	454.700	0.007	-0.006	0.091	0.088	12.410	0.001	0.006
	04/23/08	0.040	0.005	80.600	-0.001	0.000	0.020	6.600	42.000	0.023	140.000	0.004	-0.001	0.092	0.114	9.510	0.000	0.001
HW 5a (s)	10/22/07	0.020	0.020	151.700	-0.001	0.003	0.030	16.900	39.400	0.004	463.900	0.001	-0.010	0.075	0.110	12.010	0.002	0.002
	10/22/07	0.030	0.021	145.000	0.000	-0.001	0.010	4.200	54.400	0.001	146.200	0.001	-0.004	0.106	0.161	13.000	0.003	0.002
	04/23/08	0.050	0.019	116.500	0.000	0.000	0.010	5.500	59.100	0.011	137.200	0.005	0.005	0.115	0.210	10.590	-0.001	0.003
HW 7a (SUP.)	10/22/07	0.050	0.019	116.500	0.000	0.000	0.010	5.500	59.100	0.011	137.200	0.005	0.005	0.115	0.210	10.590	-0.001	0.003
	10/22/07	0.046	0.022	107.200	-0.001	-0.001	0.010	8.900	51.500	0.001	296.500	0.002	0.008	0.097	0.193	10.710	0.003	-0.001
	04/23/08	0.050	0.020	84.500	0.000	0.000	0.020	6.700	40.500	0.002	156.200	0.004	0.007	0.080	0.100	8.900	-0.001	-0.001
HW 8 (SUP.)	10/22/07	0.130	0.010	81.200	0.001	0.001	0.060	9.300	19.400	0.004	137.500	0.002	0.016	0.084	0.146	10.440	0.002	0.002
	04/23/08	-0.010	0.014	81.200	-0.001	0.000	0.010	5.100	40.300	0.001	158.900	0.002	-0.012	0.080	0.100	8.770	-0.001	-0.001
	04/23/08	0.130	0.015	88.900	0.000	-0.001	0.050	7.500	44.700	0.005	122.700	0.006	0.026	0.106	0.182	10.020	-0.001	0.000
HW 16a (s)	10/22/07	0.090	0.019	55.400	0.000	-0.001	0.040	9.500	41.200	0.177	143.000	0.003	0.022	0.070	0.184	9.760	0.003	0.002
	10/22/07	0.020	0.013	170.800	0.001	-0.002	0.010	13.100	100.500	0.004	378.200	0.005	-0.004	0.192	0.260	14.390	0.002	0.001
	04/23/08	0.070	0.012	110.000	0.002	0.015	0.170	6.400	63.700	0.026	261.200	0.006	0.006	0.134	0.177	12.180	0.002	0.025
HW 17a (s)	10/22/07	0.040	0.021	98.500	0.000	-0.002	0.020	10.700	72.900	0.000	317.400	0.001	0.006	0.145	0.200	13.620	0.002	0.001
	10/22/07	0.020	0.014	87.600	0.000	-0.005	0.010	6.800	49.300	0.001	164.200	0.001	-0.002	0.107	0.107	10.220	0.003	0.000
	04/23/08	0.010	0.016	81.500	0.001	-0.001	0.020	5.000	46.200	0.002	177.400	0.003	0.001	0.086	0.098	9.690	0.000	0.001
HW 18a (s)	10/22/07	0.040	0.011	82.700	0.001	-0.002	0.020	6.100	35.500	0.002	87.800	0.001	0.000	0.089	0.091	10.420	0.002	0.002
	10/22/07	0.040	0.017	97.300	0.001	-0.002	0.010	8.100	53.600	0.001	183.600	0.003	0.006	0.121	0.135	9.830	0.004	0.000
	04/23/08	0.070	0.016	80.200	0.002	-0.001	0.090	5.600	47.800	0.007	161.400	0.002	0.005	0.102	0.109	9.190	0.002	0.001
HW 19a (s)	10/22/07	0.090	-0.001	69.800	0.001	-0.004	0.010	5.600	41.200	0.001	125.200	-0.003	-0.008	0.079	0.093	9.770	0.002	0.001
	10/22/07	-0.030	0.019	98.200	-0.001	0.000	0.060	3.600	45.300	0.000	141.500	-0.002	-0.021	0.083	0.117	11.940	0.003	0.000
	04/23/08	-0.050	0.010	42.600	-0.002	0.001	0.000	1.300	48.900	0.306	135.200	0.017	-0.015	0.085	0.147	9.180	0.000	0.001
HW 20a (s)	10/22/07	-0.020	0.009	101.000	-0.002	-0.001	0.000	3.100	36.800	0.001	113.300	0.000	-0.010	0.067	0.111	11.920	0.002	0.001
	10/22/07	0.000	0.008	160.700	-0.001	0.000	0.010	14.500	50.400	0.001	418.500	0.003	-0.004	0.063	0.163	12.370	0.001	-0.001
	04/23/08	0.010	0.008	109.200	0.000	-0.001	0.010	6.000	35.600	0.001	230.700	0.002	-0.009	0.053	0.140	9.730	-0.003	0.000
HW 21a (SUP.)	10/22/07	0.090	0.011	160.300	0.000	-0.001	0.010	0.400	35.200	0.001	31.500	0.003	-0.005	0.048	0.057	11.050	0.001	0.006
	04/23/08	0.070	0.010	101.700	0.001	0.001	0.130	1.700	32.800	0.014	89.200	0.005	0.006	0.039	0.064	9.610	0.002	0.004
	06/28/08	0.030	0.007	95.900	0.000	-0.002	0.010	5.500	33.300	0.001	80.000	0.005	-0.003	0.045	0.059	9.970	0.001	-0.002
DITCH 1 (SUP.)	10/22/07	-0.040	0.008	107.400	0.000	-0.002	0.000	1.200	33.600	0.001	82.600	-0.001	-0.013	0.049	0.057	11.020	0.001	0.003
	10/22/07	-0.010	0.010	81.000	0.001	-0.002	0.010	7.200	49.800	0.058	155.400	0.002	-0.007	0.115	0.125	10.380	0.002	0.015
	10/22/07	-0.220	0.007	0.000	0.001	-0.007	-0.050	-3.300	0.200	0.000	0.000	-0.007	-0.027	-0.026	0.000	-0.000	-0.002	-0.004
DITCH 13 (SUP.)	10/22/07	0.050	0.009	55.700	0.000	0.001	0.100	2.100	56.900	0.010	102.100	-0.001	-0.017	0.049	0.009	4.000	0.004	0.000
	04/23/08	-0.070	0.014	48.700	0.001	-0.005	-0.010	-0.100	50.200	0.034	74.000	-0.003	-0.025	0.038	0.097	6.720	0.003	-0.004
	10/22/07	-0.050	0.010	63.100	0.000	-0.003	0.000	5.700	69.100	0.039	154.500	0.003	-0.017	0.114	0.127	10.060	0.003	-0.004
L3	04/23/08	77.50	1.60	8470.00	-1.05	0.00	368.50	4870.00	1300.00	37.35	37405.00	1.70	19.05	-5.20	15.55	47.00	-0.60	0.15
	06/28/08	25.00	4.80	7305.00	0.35	-1.10	312.50	1180.00	1200.00	32.60	35085.00	0.80	3.45	5.30	15.30	14.00	-0.10	-0.35
L2	10/22/07	0.970	-0.001	-0.100	0.004	0.003	0.150	0.600	-0.200	0.075	-0.300	0.040	0.115	1.131	0.746	25.420	-0.664	0.028
	06/28/08	0.950	0.013	-0.100	0.004	0.255	0.150	0.000	-0.200	0.036	-0.300	0.033	0.120	1.097	0.800	29.000	-0.065	0.403

METAL ANALYSES FOR SEDIMENT SAMPLES AT THE EVERSON SITE, RICHLAND, CO. (TOTAL DIGESTIBLE METALS IN SEDIMENT)

(TOTAL RECOVERABLE)

LOC.	DEPTH	ppm Al	ppm Ca	ppm Fe	ppm K	ppm Na	ppm Si	ppm As	ppm Cd	ppm Cu	ppm Hg	ppm Mn	ppm Ni	ppm Pb	ppm Sb	ppm P	ppm Ti	ppm Zn
PITRUD 1	6-8 ft.	3.773	6.379	2.101	1.960	1.003	30.330	-26.000	1.800	18.000	1.447	397.200	10.000	30.000	42.000	0.045	1293.900	49.000
PITRUD 2	6-8 ft.	3.402	7.706	2.199	1.572	2.103	25.646	-16.016	1.001	20.270	1.733	411.512	106.100	44.000	55.055	0.044	1425.323	50.051
HW 3	5-7 ft.	4.121	3.037	2.501	2.191	1.630	32.211	-30.076	1.494	14.940	0.942	434.661	84.661	21.912	21.912	0.040	1531.765	56.574
HW 4	6-12 ft.	4.310	4.519	2.902	2.077	2.258	29.690	-24.720	1.000	40.752	1.607	459.030	156.701	16.815	45.500	0.069	2208.124	91.900
HW 7	2-7 ft.	3.360	2.674	3.504	2.025	0.026	25.935	-36.378	1.191	15.285	0.920	632.139	31.762	-30.760	-14.800	0.050	1507.295	57.965
HW 7 (SUP.)	2-7 ft.	3.779	2.910	4.106	2.733	0.072	32.010	-17.066	1.290	17.469	1.050	740.844	32.754	-13.094	29.777	0.064	1777.560	67.393
FL. UNION	20 ft.	3.755	2.823	1.020	1.042	0.737	30.523	-21.901	0.096	13.041	3.217	435.341	173.221	-20.706	97.561	0.059	1445.396	46.391

(NEGATIVE INDICATES LESS THAN DETECTION LIMIT).

APPENDIX L
BASIC ION AND METAL ANALYSES
(MBMG LAB)

WATER ANALYSES RESULTS FOR THE IVERSON SITE, RICHLAND, CO. (DISSOLVED)

PARAMETER	DATE	LOCATION	MW 2	MW 21s	MW 3	MW 3s	MW 4s	MW 4b	MW 4d	MW 5s	MW 6	MW 7s	MW 9	MW 13	MW 16s	MW 17s	MW 18s	MW 19s	MW 20s	L3 DITCH 13	DITCH 1	
ppm Ca	07/22/87	C-1	145	111	357	226	80.7	84.9	119	61	71.5	185	91.6	94.6	98.9	167	66.8	8350	62.2	55.3	66.8	
	10/22/87			102	95.2	172	83.2	86.4	109	62.4	88	108	7310	53.2								
	04/23/89			105		147		98.6	60.7	73	110											
	06/28/89																					
ppm Mg	07/22/87	C-1	31.2	34	90.7	50.6	54.9	45.8	53.6	45.2	45.6	112	50.3	53.2	44.4	50.2	73.1	1270	34.7	72.4	73.1	
	10/22/87			32.4	37.4	44.5	42.5	55.4	53.2	47.4	45.2	65.5	42.3	44.3	45.2	35.3		1260	40.7			
	04/23/89			37.2		39.7	43	46.7	43.3	47.5		77.6	38.5	43.9	40			1300	33.2			
	06/28/89																					
ppm Na	07/22/87	C-1	63.5	85.2	1436	681	192	112	196	156	144	370	162	177	135	437	163	8170	97.2	161	163	
	10/22/87			91.8	105	462	143	145	211	147	257	161	149	131	232	88600	121					
	04/23/89			89.1		324	476	130	146	161	327	89.6	134	116				73500	81.6			
	06/28/89																					
ppm K	07/22/87	C-1	3.8	4.5	25.4	26.4	6.9	5	8.2	5.7	6.5	13.5	6.7	7.7	6.7	15	10.2	1040	5.2	10	10.2	
	10/22/87			3.3	4.5	18.7	0.6	6.4	8.8	5.2		8.6	5.3	5.5	5.4	8.8						
	04/23/89			5		14.4	19.2	5.6	7.3	7.1		11	5.9	7.3	6.7							
	06/28/89																					
ppm Fe	07/22/87	0.063	<.002	<.002	0.004	0.002	<.002	<.002	0.005	<.002	0.007	<.002	<.002	<.002	<.002	<.002	<.002	<.002	399	0.072	<.002	0.002
	10/22/87			0.012	0.11	<.002	0.014	0.016	0.066	0.35		0.023	0.051	0.044	0.28	0.099			358	0.04		
	04/23/89			<.002	<.002	<.002	<.002	<.002	<.001	0.005		<.002	<.002	<.002	<.002	<.002			379	0.006		
	06/28/89																					
ppm Mn	07/22/87	<.001	<.001	5.01	0.29	0.053	0.001	0.003	0.5	0.25		0.001	<.001	<.001	<.001	<.001	<.001	42	0.022	0.039	0.058	
	10/22/87			0.4	0.085	0.022	0.015	0.015	0.21	0.065		0.02	0.016	0.009	0.065	0.006			35.9	0.017		
	04/23/89			0.28	0.001	0.001	0.001	0.001	0.19	0.001		<.001	0.001	0.001	0.001	0.001			30.9	0.037		
	06/28/89																					

(DISSOLVED)

PARAMETER	DATE	0A/0C	MW 2	MW 21s	MW 3	MW 3s	MW 4s	MW 4p	MW 4d	MW 5s	MW 6	MW 7s	MW 8	MW 13	MW 16s	MW 17s	MW 18s	MW 19s	MW 20s	L3 DITCH 13	DITCH 1	
pea F	07/22/87	0.1	0.2	0.4	0.3	0.3	0.8	0.8	0.8	0.4	0.6	0.6	0.7	0.7	0.4	0.6	0.6	0.6	0.6	10	0.4	
	10/22/87		0.6	0.6	0.6					0.4	0.7	0.7			0.4	0.6	0.6	0.6	0.6		0.6	
	04/23/88		0.7	0.6	0.6	0.8				0.5	0.3	1			0.5	0.4	0.8	0.8	0.4	2	0.3	
	06/28/88			0.5	0.5	0.5				0.5	0.8	0.8	1.1		0.5	0.7	0.7	0.7	0.7	430	0.2	
poa Br	07/22/87	0.1	0.1	0.1	2.5	0.1	0.4	0.4	0.4	0.1	0.1	0.5	0.1	0.2	0.4	0.2	0.3	0.1	0.7	110	0.1	
	10/22/87		0.1	0.1	0.6					0.1	0.4	0.4			0.4	0.2	0.3	0.1	0.7		0.1	
	04/23/89		0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.2	0.1	
	06/28/89		0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
poa P	07/22/87	0.1	0.1	0.1	0.2	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
	10/22/87		0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
	04/23/89		0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
	06/28/89		0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
OS	07/22/87	0.3	693.72	5113.98	2829.98	989.16	744.43	1160.22	786.92	1950.49	915.58	982.82	843.58	1837.07	1785	898	846	827	1146	624.3	955.73	966.39
	10/22/87		688.85	1086.67	1899.51	825	882.77	1095.28	800	954	967	800			1459.43	596.69	757.88	782.35		715		
	04/23/89		766.62	760				823.84	789.89											513.15		
	06/28/89		692.07	1435.81	1935.42																	
SUM DISS	07/22/87	0.4	870.8	5383.92	2928.83	1223.36	973.27	1341.73	1022.35	2260	1155.58	1225.36	1080.03	2100.92	1580	1158	1102	1077	1407	762.31	1209.93	1213.49
	10/22/87		911.09	1332.55	2172.99	1079	1293.42	1337.82	1047	1199	1164	1047			1717.7	778.34	959.97	987.95		866		
	04/23/89		928.9	99E				1051.15	1021.77											645.89		
	06/28/89		885.9	1659.57	2193.18																	
LAB SC	07/22/87	2.9	1178.9	8445.8	5059.1	1602.2	1187.1	1860.9	1276.1	1277.2	3534	1555	1681	1421	3362	1794	100900	1584	113502	1037.3	1573	1584
	10/22/87		1139	1921	3194			1942												3362		
	04/23/88		1125	1165	1369			1418	1244	2192	1440	1345	1242	1794	100900	1544				100900		
	06/28/88		1067	1067	2445	3313	1432	1317	1213	2413	934.9	1179	1234									

(Discovered)

PARAMETER	DATE	BR/OC	MW 2	MW 21s	MW 3	MW 3s	MW 4s	MW 4e	MW 4d	MW 5s	MW 6	MW 7s	MW 8	MW 13	MW 16s	MW 17s	MW 18s	MW 19s	MW 20s	LY DITCH 13	DITCH 1	
ppm S102	07/22/87	0.1	20.5	19	19	31.1	19.9				23.8	24.3	21.5	21.3						8.3	14.4	
	10/22/87			25.4	28.5	27.9				29.6	23.8	23.9	20.4		32.1	22.8	21.5	26	26.6		22.8	22.3
	04/23/88			22.3	21.8		22.1			23.9	21.8	21.8	20.4		28	21	19.7	22.6	23.8		7.5	9.2
	06/28/88			23.8		29.3				27.3	23.5	23.5	22.6		31.9	24	21.8	27.6			1.3	15.7
ppm HED3	07/22/87	0.2	349		532		589	501			451	476	464	453						220.8	272	
	10/22/87			438	484		539			500	478	478	486		610	473	478	466	520		501	487
	04/23/88			438	470			501		483	506	506	486		583	512	505	492	514		332	299
	06/28/88			382		441	508			475	448	448	457		509	358	418	405			268.4	261.6
ppm C03	07/22/87	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
	10/22/87			0	0					0	0	0	0		0	0	0	0	0		0	14.6
	04/23/88			0	0			0		0	0	0	0		0	0	0	0	0		0	0
	06/28/88			0	0					0	0	0	0		0	0	0	0	0		0	0
ppm C1	07/22/87	0.1	31.9		2660		1060	117			37.4	217	31	51.8						14500	25.9	
	10/22/87			19.7	221		678			186	224	224	31.4		626	119	146	87.9	646		33.8	30.9
	04/23/88			21.1	24.4			28.2		130	70.7	70.7	31.4		268	71.3	44.4	49.4	220		156000	21.4
	06/28/88			23.2		411	680			83.7	60.6	60.6	29.2		393	20.8	30.2	56.1			130000	17.7
ppm S04	07/22/87	0.1	203		256		253	250			210	243	236	224						690	2249	
	10/22/87			187	208		227			213	221	221	244		295	227	244	211	234		340	342
	04/23/88			213	233			251		238	267	267	245		247	265	255	240	262		1110	305
	06/28/88			209		243	260			231	234	234	235		259	172	240	218			223	174
ppm N	07/22/87	0.02	22.7		2.51		9.14	0.11			2.77	4.03	0.75	0.23						25	1.22	
	10/22/87			5.89	3.85		3.21			9.42	2.22	2.22	3.53		16	2.59	2.76	3.53	4.52		33	2.93
	04/23/88			4.34	1.55			0.97		8.3	1.24	1.24	0.85		5.25	1.25	1.15	1.78	2.3		0.04	0.67
	06/28/88			11		10.6	3.28			4.98	2.75	2.75	0.39		10.1	2.24	1.07	7.75			36	2.35

PARAMETER	DATE	04/08	MW 2	MW 21s	MW 3	MW 3s	MW 4s	MW 4a	MW 4d	MW 5s	MW 6	MW 7s	MW 8	MW 13	MW 16s	MW 17s	MW 18s	MW 19s	MW 20s	L3 DITCH 13	DITCH 1
LAP pH	07/22/87	5.05	7.52	7.48	7.39	7.01	7.17	7.69	7.69	7.5	7.74	7.56	7.71	7.71	7.64	7.68	7.32	7.34	7.32	5.25	8.06
	10/22/87			7.29	7.51	7.01	7.01			7.71		7.47	7.43	7.71	7.9	7.7	7.71	7.52	7.48		8.35
	04/23/88			7.86		7.48	7.6	7.59		7.85		7.62	7.89		7.89	7.83	8.07	7.91			8.35
	06/28/88																			4.98	7.92
HARDNESS	07/22/87	0	490.46	1264.75	1264.75	772.59	772.59	427.47	427.47	400.5	517.76	339.36	339.36	28973.79	298.13	298.13	298.13	298.13	298.13	28973.79	298.13
	10/22/87			417.11	429.51	612.64	612.64			598.4	491.14	491.14	429.7	623.62	922.93	435.76	455.18	429.7	623.62	461.05	467.67
	04/23/88			388.05	394	383	383			526	389	351	351		564	370	372	406	415	26036	323
	06/28/88			415.3	526.34	613.96	613.96			489.36	390.21	347.07	347.07		565.6	324.76	362.97	439.31	23603.87	268.49	
ALKALIN	07/22/87	0.16	286.23	436.33	436.33	463.08	463.08	410.9	410.9	369.89	390.4	380.55	380.55	181.09	223.08	223.08	223.08	223.08	223.08	181.09	223.08
	10/22/87			359.23	396.96	442.07	442.07			410	392.04	392.04	382.19	426.48	500.3	387.94	392.04	382.19	426.48	410.9	423.77
	04/23/88			359.23	385	411	411			396	415	399	399		478	420	414	404	422	272.29	245
	06/28/88			313.3	361.69	416.64	416.64			389.58	367.43	374.81	374.81		417.46	293.62	342.83	332.16	220.13	214.55	
ION BAL	07/22/87	0.055	-0.640	0.498	0.498	-0.413	-0.413	-0.877	-0.877	0.145	560	-0.758	-0.758	-0.517	0.117	0.361	0.749	0.502	0.094	0.604	0.003
	10/22/87			-0.315	0.091	0.204	0.204			0.367	-0.382	-0.382	-0.382		0.117	0.361	0.749	0.502	0.094	0.035	-0.003
	04/23/88			2.525	2.663	0.985	0.985			1.63	2.934	1.675	1.675		-0.049	4.571	2.919	2.098	4.253	0.032	
	06/28/88			-0.78	-0.623	-0.592	-0.592			-0.732	-0.56	-2.747	-2.747		-0.51	-1.198	-1.551	-2.13	-2.13	-2.179	

METAL ANALYSES OF WATER SAMPLES AT THE IVERSON SITE, RICHLAND, CO. (DISSOLVED)

PARAMETER	DATE	LOCATION						L3
		MW 21s	MW 3	MW 3s	MW 4s	MW 16s	MW 20s	
ppb Al	07/21/87		<30					
	10/22/87		<30				<30	
	04/23/88		<30					80
	06/28/88	<30		<30	<30	<30		150
ppb Ag	07/21/87		<2					
	10/22/87		<2				<2	
	04/23/88		<2					44
	06/28/88	<2		<2	<2	<2		94
ppb B	07/21/87		1030					
	10/22/87		220				570	
	04/23/88		140					38000
	06/28/88	120		400	820	600		38000
ppb Cd	07/21/87		<2					
	10/22/87		<2				<2	
	04/23/88		<2					16
	06/28/88	<2		<2	<2	<2		30
ppb Cr	07/21/87		<2					
	10/22/87		<2				<2	
	04/23/88		<2					14
	06/28/88	<2	<2		<2	<2		12
ppb Cu	07/21/87		<2					
	10/22/87		4				<2	
	04/23/88		<2					1920
	06/28/88	<2		<2	11	<2		890
ppb Li	07/21/87		370					
	10/22/87		71				160	
	04/23/88		20					13100
	06/28/88	20		130	190	100		10000
ppb Mo	07/21/87		<20					
	10/22/87		<20				<20	
	04/23/88		<20					50
	06/28/88	<20		<20	<20	<20		20
ppb Ba	07/21/87		530					
	10/22/87						92	
	04/23/88		60					320
	06/28/88			110	100			270
ppb Pb	07/21/87		<40				<40	
	10/22/87							
	04/23/88		<40					350
	06/28/88			<40	<40			270

CONT.

PARAMETER	DATE	MW 21s	MW 3	MW 3s	MW 4s	MW 16s	MW 20s	L3
ppb Ni	07/21/87		<10					
	10/22/87		<10				<10	
	04/23/88		<10					160
	06/28/88	<10		<10	<10	<10		160
ppb Sr	07/21/87		7300					
	10/22/87		1020					
	04/23/88		990				2510	241000
	06/28/88	630		2660	2780	940		207000
ppb Ti	07/21/87		29					
	10/22/87		3					
	04/23/88		<1				5	<1
	06/28/88	4		6	8	5		<1
ppb V	07/21/87		<1					
	10/22/87		<1					
	04/23/88		<1				<1	51
	06/28/88	<1		<1	<1	<1		15
ppb Zn	07/21/87		14					
	10/22/87		4					
	04/23/88		<3				<3	880
	06/28/88	<3		<4	3	<3		430
ppb Zr	07/21/87		<4					
	10/22/87		<3					
	04/23/88		<4				<4	66
	06/28/88	<4		<4	<4	<4		66
ppb As	07/21/87		0.5					
	10/22/87							
	04/23/88		0.7				0.5	0.3
	06/28/88			<.1				2
ppb Se	07/21/87							
	10/22/87							
	04/23/88		1.2				1	0.6
	06/28/88							2.2
ppb Be	07/21/87		21					
	10/22/87							
	04/23/88		<1					26
	06/28/88			1				21

APPENDIX M
CLAY MINERALOGY

Iverson Site Pit Mud:

05/25/88

glycol

Full Scale 1029

.1 step

evaporated

cps SF= 1.75

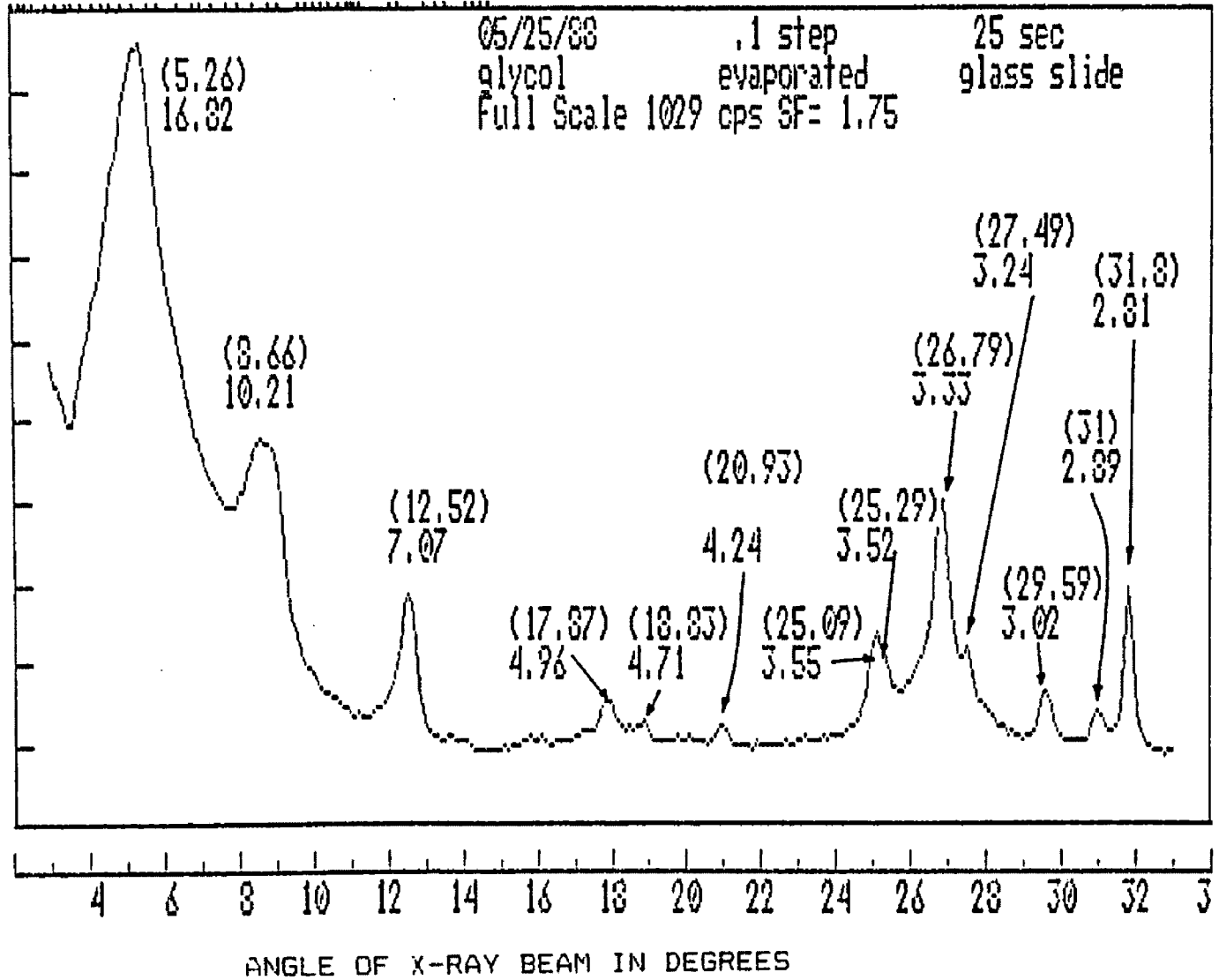
25 sec

glass slide

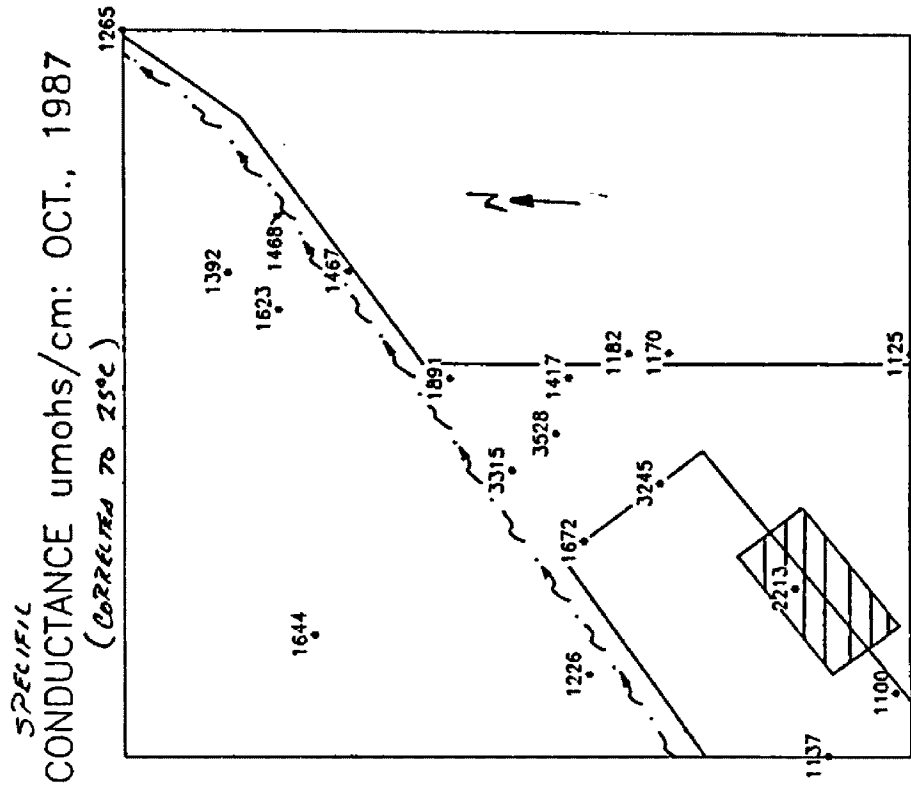
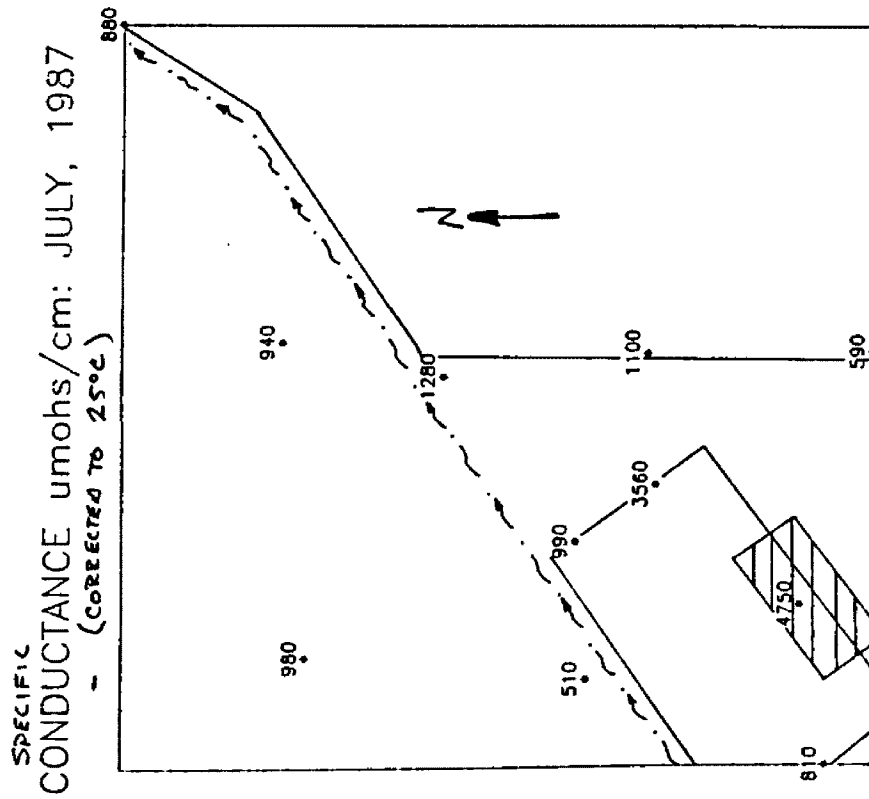
ANGSTROMS - INTERPRETATION

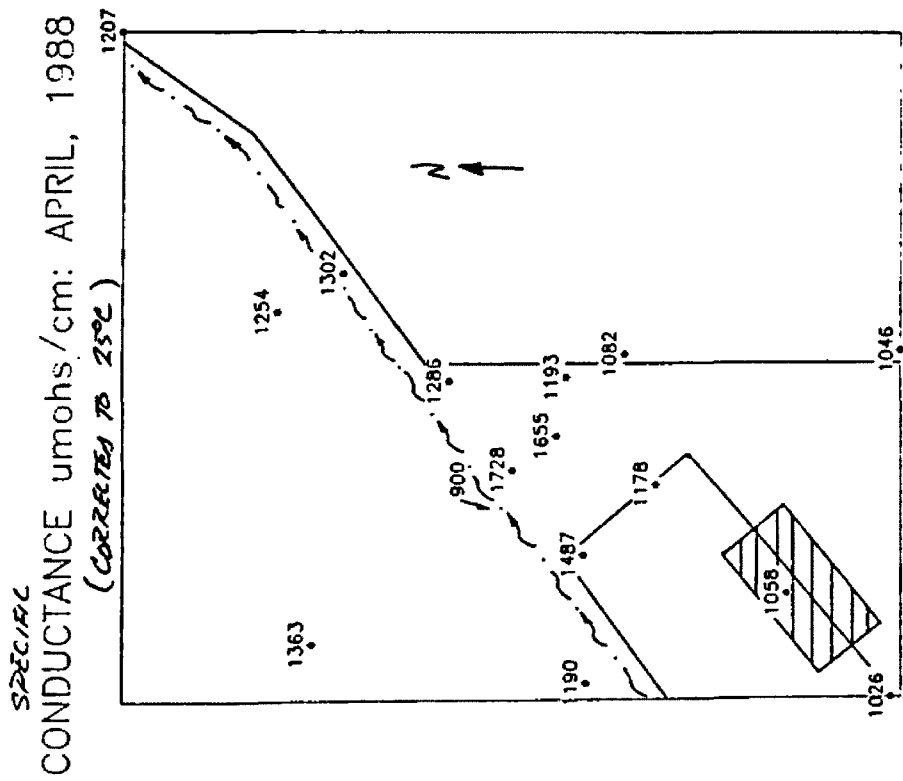
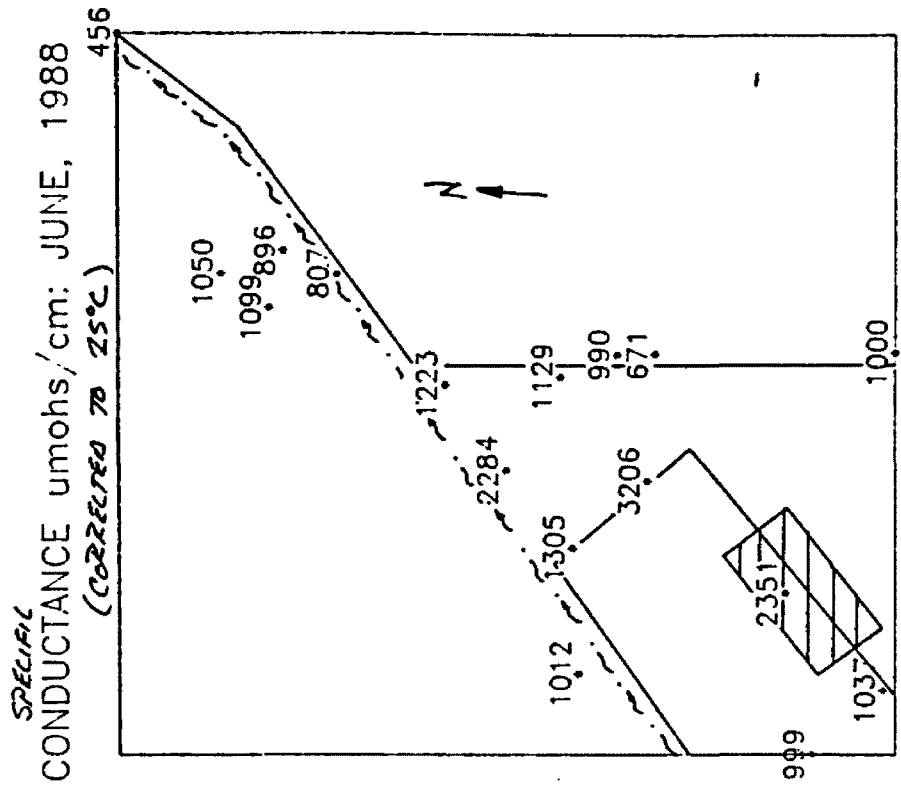
- 16.82 - SMECTITE 001
- 10.21 - ILLITE 001
- 7.07 - KAOLINITE 001
OR CHLORITE 002
- 4.96 - ILLITE 002(?)
- 4.71 - ILLITE 002(?)
- 4.24 - SMECTITE 002(?)
- 3.55 - CHLORITE 003
- 3.52 - KAOLINITE 002
- 3.33 - QUARTZ
- 3.24 - K-SPAR(?)
- 3.02 - CALCITE
- 2.89 - DOLOMITE
- 2.81 - HALITE

RELATIVE STRENGTH

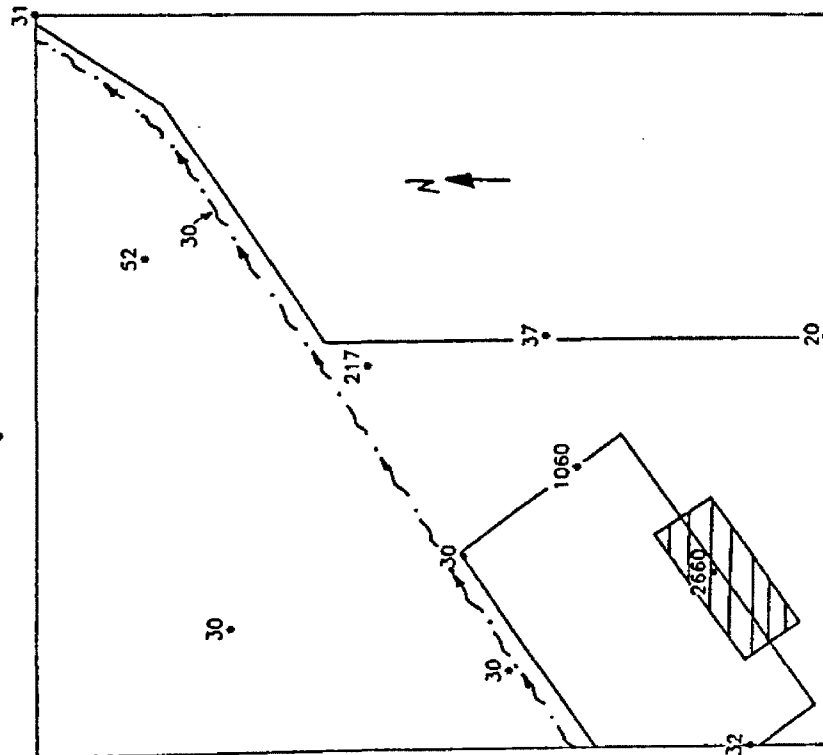


APPENDIX N
CHLORIDE DATA AND SC DATA/PLOTS

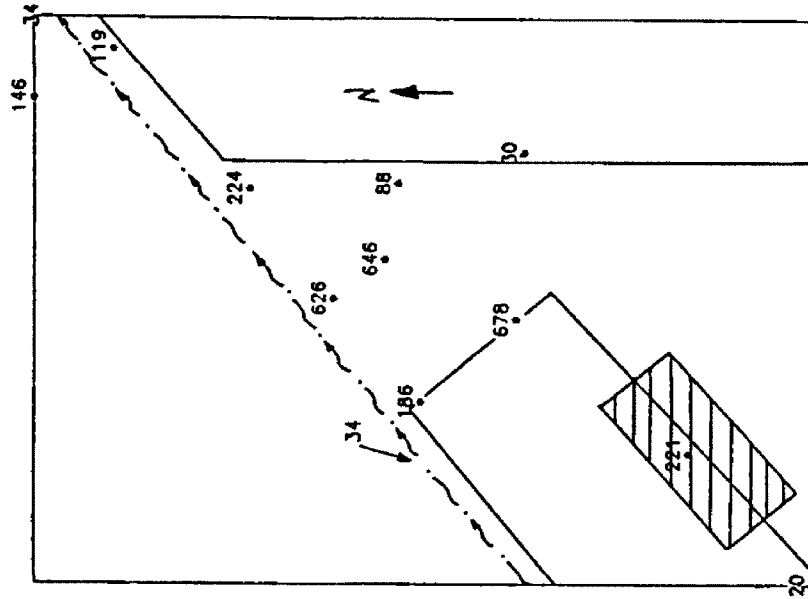


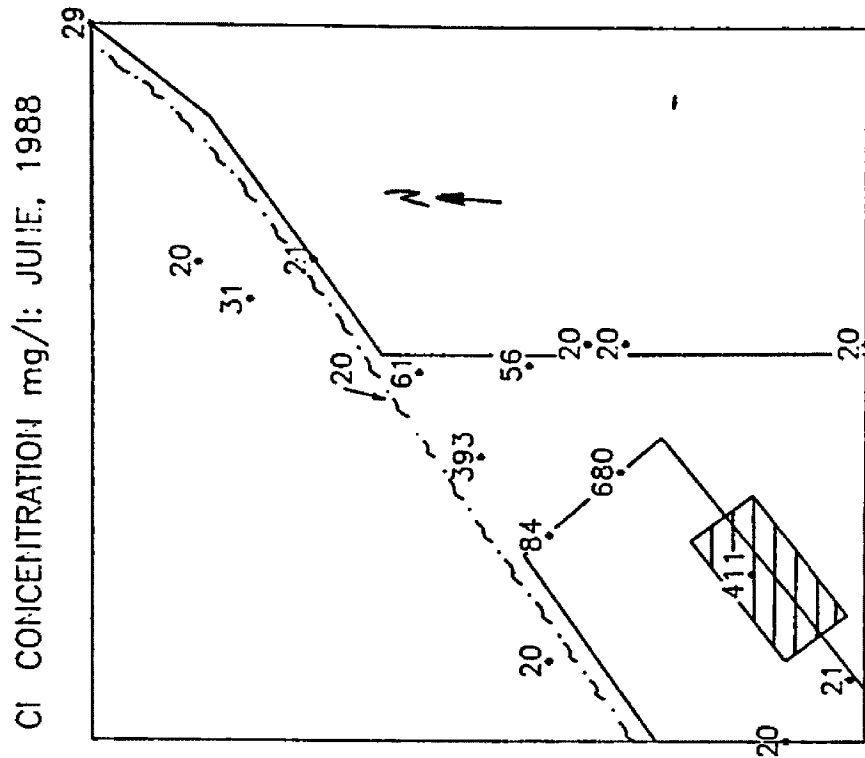
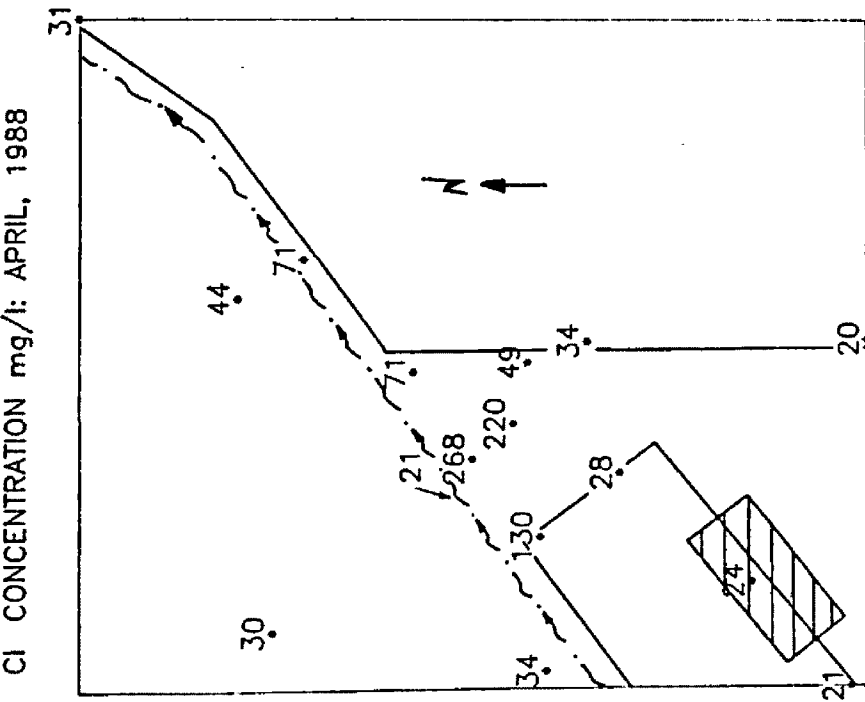


Cl CONCENTRATION mg/l: JULY, 1987

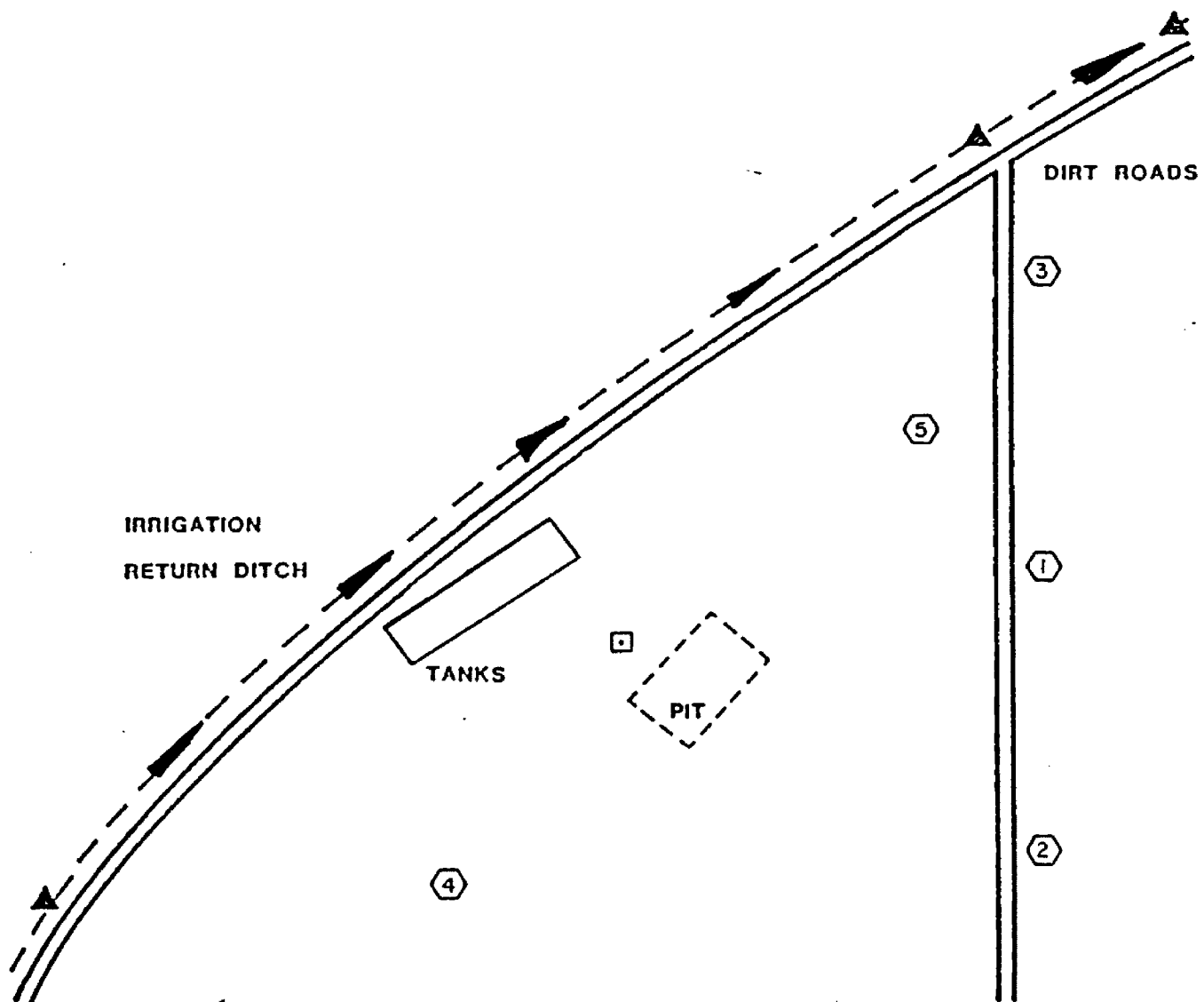


Cl CONCENTRATION mg/l: OCT., 1987





APPENDIX O
YELLOWSTONE RIVER CHEMISTRY AND DEWEY'S DATA



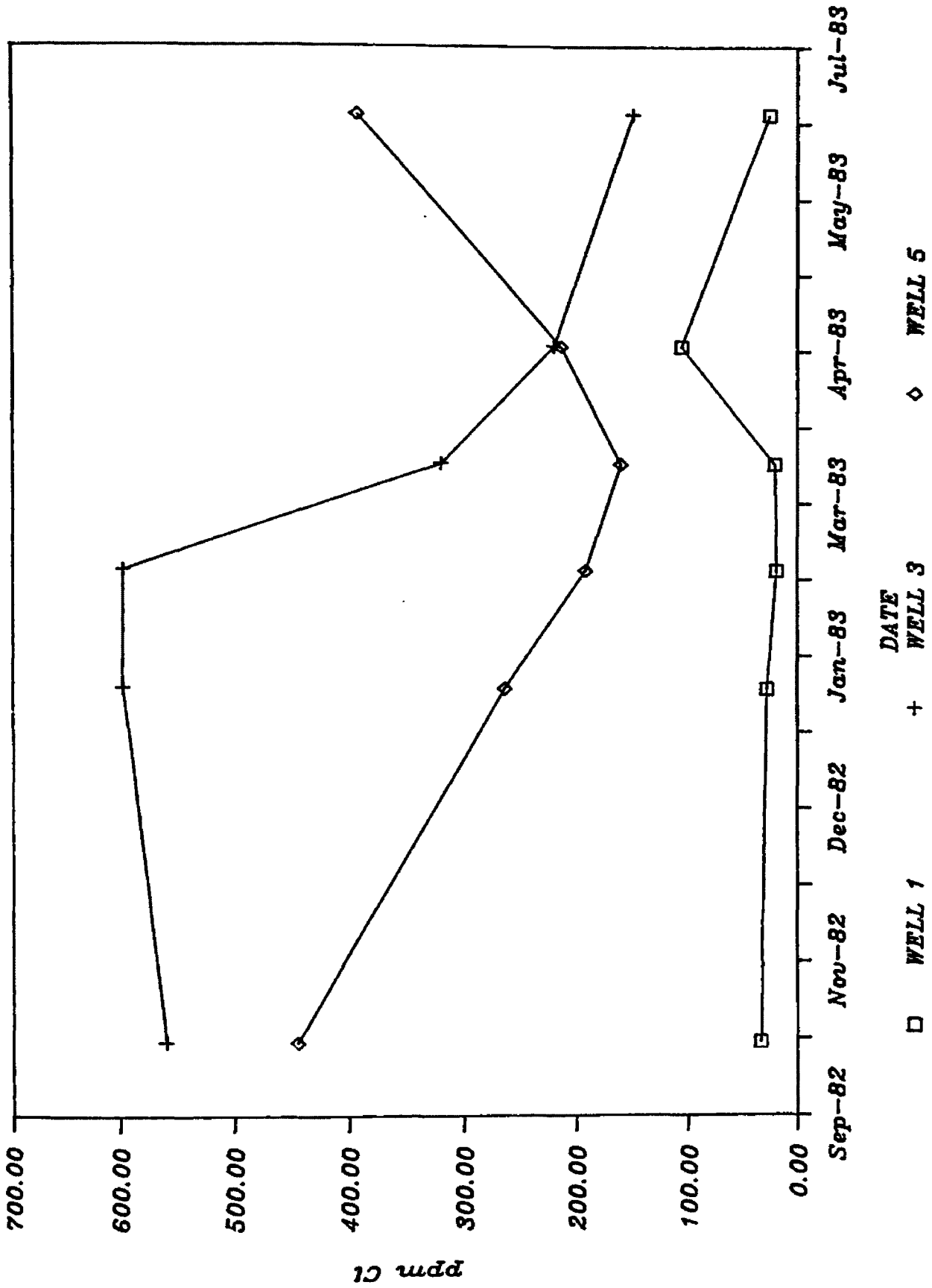
KEY

0	100 ft	N
0	30 m	
□	OIL WELL	
①	MONITORING WELL	
▲	SURFACE SAMPLE SITE	

**DEWEY'S
MONITORING WELL
SITES**

DEWEY'S CHLORIDE CONC. IN ppm

IVERSON SITE



DATA PRESENTED BY DEWEY CHEMICAL DATA FOR HER WELLS FROM 1982-1983
 (NOTE MAP IN THIS APPENDIX FOR WELL LOCATION)

CHLORIDE CONCENTRATION IN ppm

	10/15/82	01/15/83	02/15/83	03/15/83	04/15/83	06/15/83
WELL 3	560	600	600	321	220	150
WELL 5	445	264	192	161	214	395
WELL 1	33.5	27.8	19.1	20.7	106	25

SULFATE CONCENTRATION IN ppm

WELL 3	265	252	260	259	260	264
WELL 5	234	223	231	240	249	247
WELL 1	217	217	225	225	240	237

SPECIFIC

CONDUCTANCE IN umhos/cm 25 C

	08/18/82	08/24/82	09/01/82	09/04/82	09/09/82	10/28/82	01/30/83
WELL 3		2219	1790	1862	1854	2675	1522
WELL 5				2085	3382	2339	2353
WELL 1	1125	1153	987	1128	1543	1231	1213

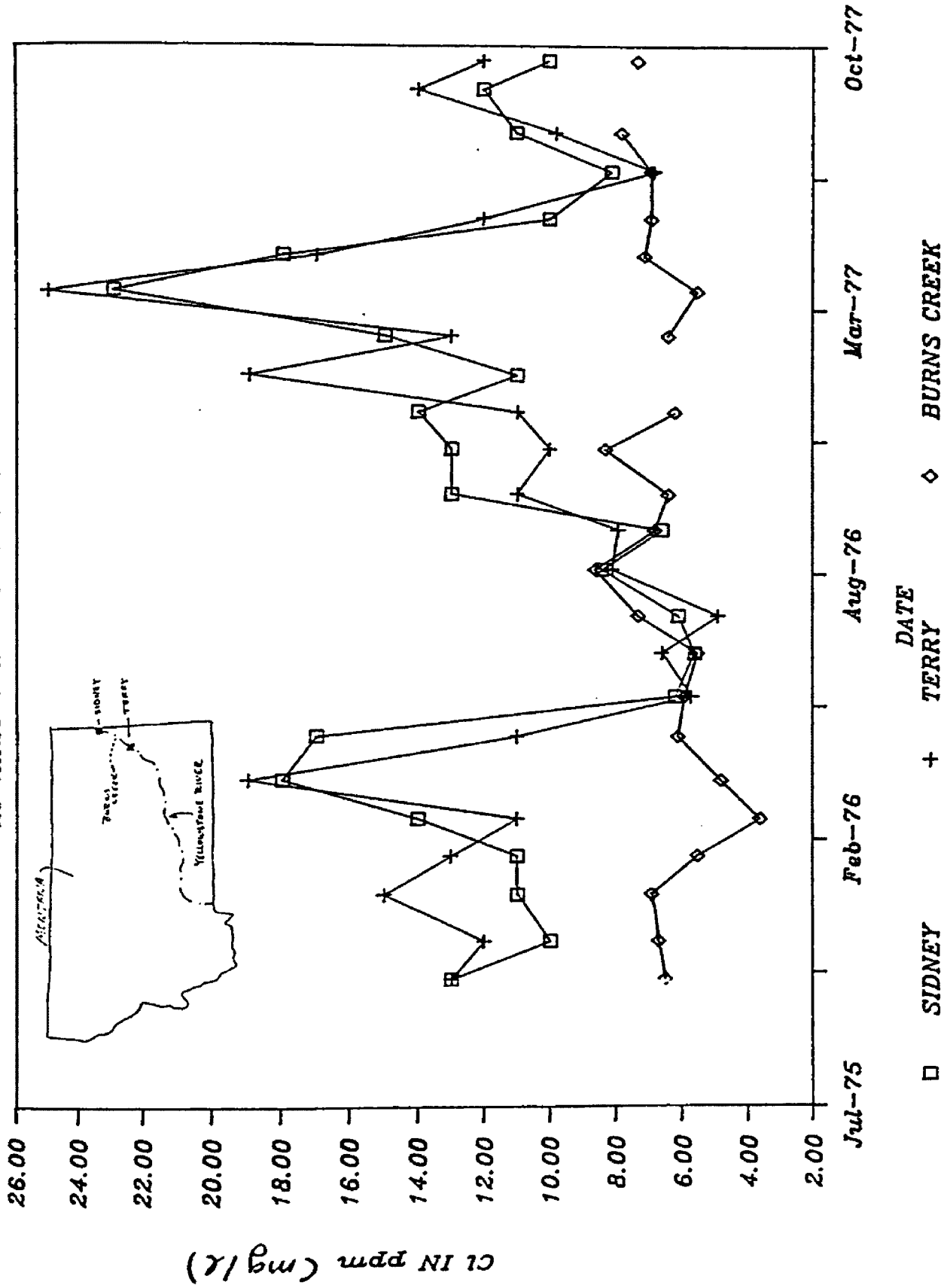
(ppm) CHEMICAL PARAMETERS AUGUST, 1982 (DISSOLVED IONS)

PARAMETER	WELL 1	WELL 2	WELL 3	DITCH BL
Ca	85.3	93.1	124.8	56.9
Mg	38.1	31.6	70	35.5
Na	114	102.4	285.7	84.2
K	6.8	6.4	11.4	6.1
HCO3	436.8	424.6	510	305
CO3	0	0	0	0
Cl	14.5	13.5	368	17.5
SO4	230	195	250	190
pH	7.77	7.69	7.78	8.21
TDS	925	866.6	1619	695.2
CONDUCT.	1139	1085	2312	911
HARDNESS	370	363	600	288
ALKALIN.	358	348	418	250
SAR	2.6	2.3	5.1	2.2

(ppm) CHEMICAL PARAMETERS FOR SEPT. 1982 (DISSOLVED IONS)

PARAMETER	WELL 1	WELL 2	WELL 3	WELL 5	DITCH BL
Ca	89.2	67.2	89.6	156.8	58.9
Mg	35.04	40.8	49.5	60.1	29.5
Na	113.6	111.2	241.9	261	81
K	7.4	8	10.1	13.4	5.8
HCO3	392.8	395.3	475.8	283.1	283
CO3	0	0	0	0	0
Cl	29	14.8	200	450	17.5
SO4	225	235	250	225	195
pH	7.81	7.95	8.04	7.78	8.06
TDS	892.4	872.3	1316.9	1549.4	670.7
CONDUCT.	1190	1136	1825	2357	882
HARDNESS	368	336	428	639	269
ALKALIN.	322	324	390	314	232
SAR	2.6	2.6	5.1	4.5	2.2

CHLORIDE IN ppm IN SURFACE WATER AT VARIOUS LOCATIONS IN E. MT.



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REGIONAL WATER QUALITY (USGS)

CHLORIDE (ppm)

LOCATION	DATE	Oct-75	Nov-75	Dec-75	Jan-76	Feb-76	Mar-76	Apr-76	May-76	Jun-76	Jul-76	Aug-76	Sep-76	Oct-76	Nov-76	Dec-76	Jan-77	Feb-77	Mar-77	Apr-77	May-77	Jun-77	Jul-77	Aug-77	Sep-77
YS/SIDNEY	13	11	14	18	17	14	18	17	14	14	14	14	14	14	14	14	14	14	14	14	14	14	14	14	14
YS/TERRY	13	15	15	19	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
BURNS CK	6.5	6.7	5.5	4.6	6.1	5.9	5.5	7.3	8.6	6.8	6.4	8.3	6.2	6.4	5.5	7.1	6.9	6.9	7.6	7.6	7.6	7.6	7.6	7.6	7.6

SODIUM (ppm)

YS/SIDNEY	66	63	69	80	80	72	80	80	80	80	80	80	80	80	80	80	80	80	80	80	80	80	80	80	80
YS/TERRY	70	62	75	68	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63	63
BURNS CK	200	200	190	266	210	230	200	230	220	220	200	279	190	190	160	240	240	240	240	240	240	240	240	240	240

SULFATE (ppm)

YS/SIDNEY	190	200	210	230	240	230	230	240	240	240	240	240	240	240	240	240	240	240	240	240	240	240	240	240	240
YS/TERRY	190	206	246	286	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220	220
BURNS CK	440	460	430	460	540	390	400	430	490	450	400	530	390	390	350	640	520	510	560	560	560	560	560	560	560

CALCIUM (ppm)

YS/SIDNEY	57	59	69	66	63	55	63	66	66	66	66	66	66	66	66	66	66	66	66	66	66	66	66	66	66
YS/TERRY	55	63	75	66	66	60	66	57	57	57	57	57	57	57	57	57	57	57	57	57	57	57	57	57	57
BURNS CK	53	70	71	52	79	68	58	44	45	52	66	90	70	70	57	77	54	51	54	54	54	54	54	54	54

POTASSIUM (ppm)

YS/SIDNEY	3.9	3.5	3.4	3.9	3.9	3.6	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9	3.9
YS/TERRY	3.7	3.4	3.6	4.4	3.7	3.2	4.4	3.7	2.4	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3	2.3
BURNS CK	8.3	7.5	8.1	7.9	7.9	8.2	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1	9.1

MAGNESIUM (ppm)

YS/SIDNEY	22	22	24	26	27	23	26	27	27	27	27	27	27	27	27	27	27	27	27	27	27	27	27	27	27
YS/TERRY	21	22	27	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24	24
BURNS CK	82	77	77	79	53	79	82	79	77	79	75	100	76	76	61	59	94	81	83	82	82	82	82	82	82

SOLIDS (ppm)

YS/SIDNEY	476	468	506	574	575	523	574	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575
YS/TERRY	476	468	506	574	575	523	574	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575
BURNS CK	476	468	506	574	575	523	574	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575	575

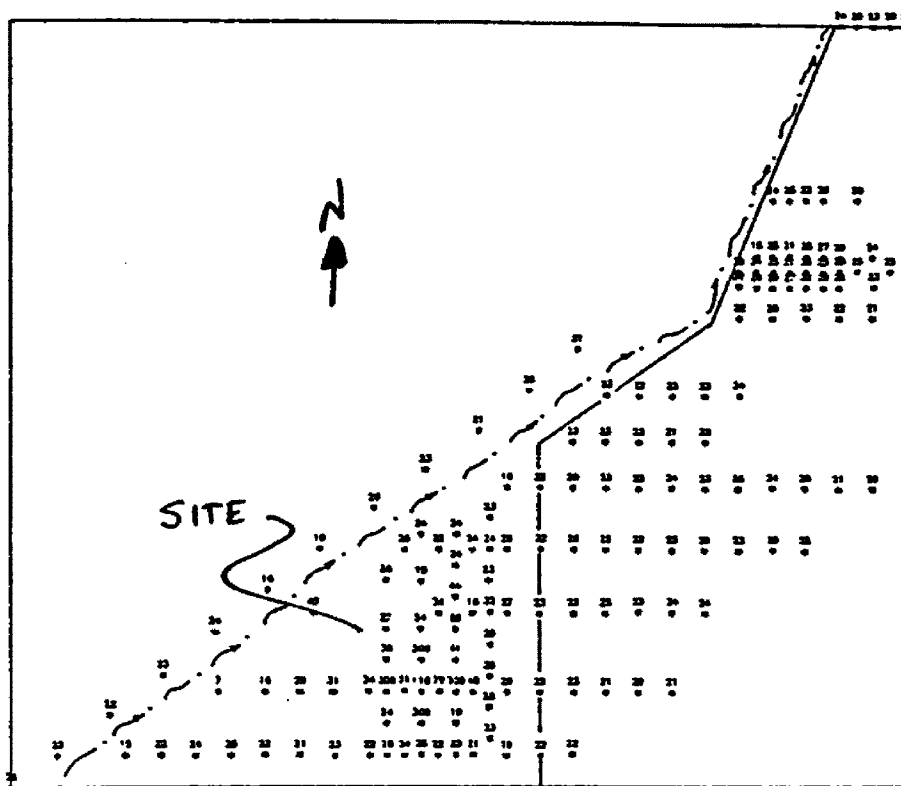
STREAM FLOW IN CFS

YS/SIDNEY	11100	8600	12600	10800	11000	13200	12100	37600	33300	13800	4440	9300	9590	7600	7600	8200	9800	7400	6700	6300	12200	10500	3830	4000	5170
YS/TERRY	9776	16600	6700	9730	10600	10600	15700	32500	42100	26500	9590	9150	10000	9360	8200	3930	8230	8230	5960	6710	11100	15800	5350	4100	5920
BURNS CK	1.7	2.3	3	2.1	63	17	6.3	6.9	4.6	0.77	0.78	0.89	18	1	1.3	1.9	3.5	3.5	2.1	1.1	0.64	0.32	0.64	0.32	0.64

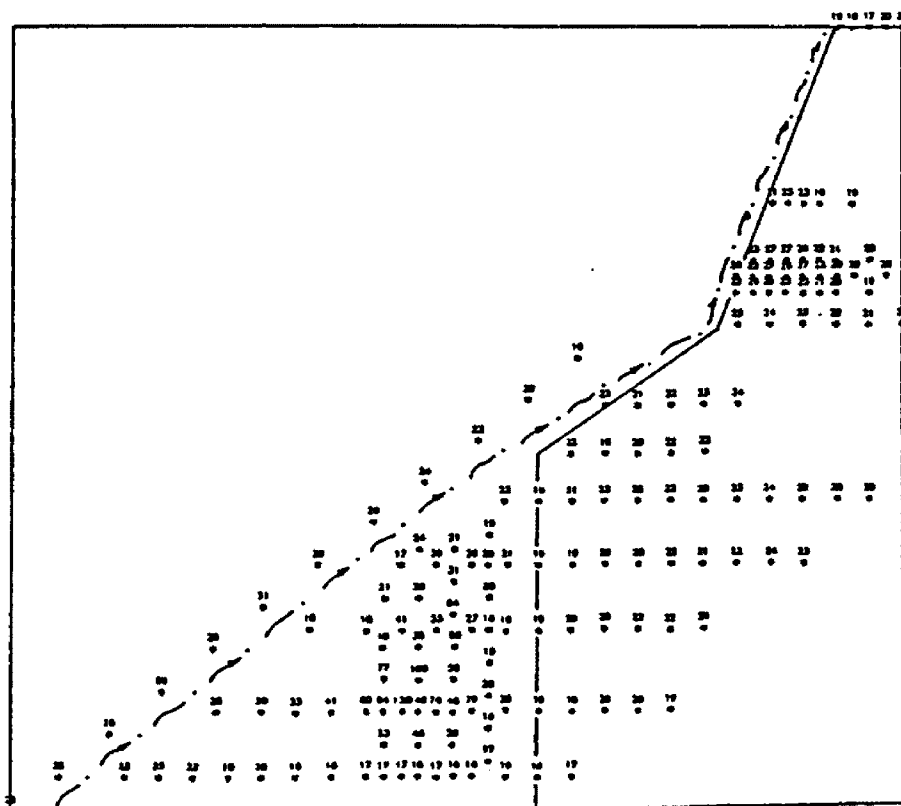
APPENDIX P

EM 10,20, AND 40 m DATA/PLOTS

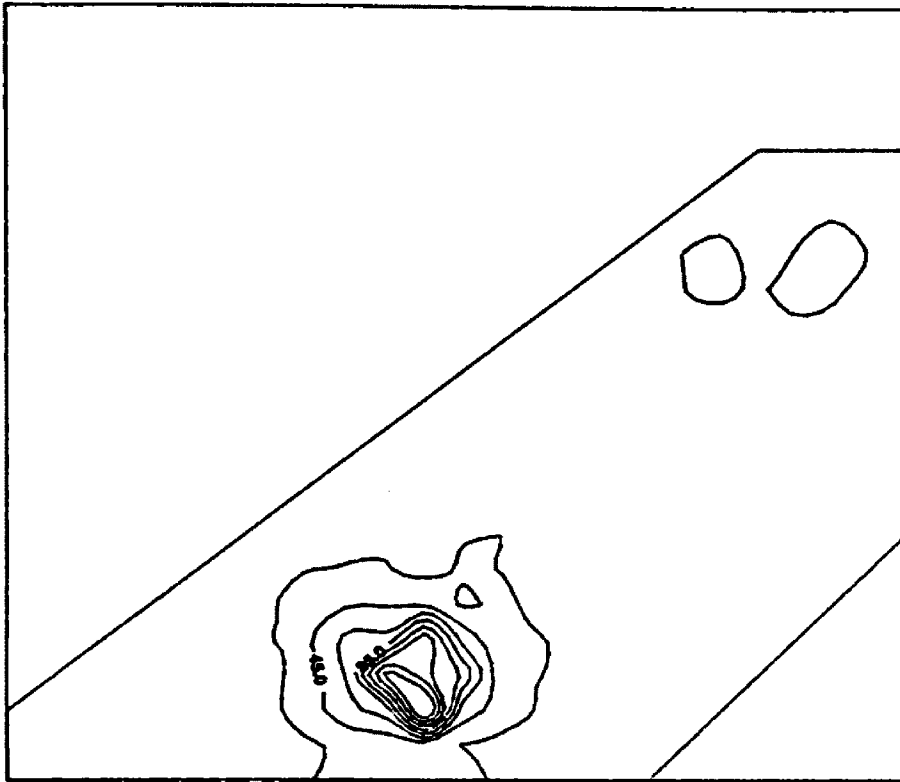
EM 10H DATA PLOT



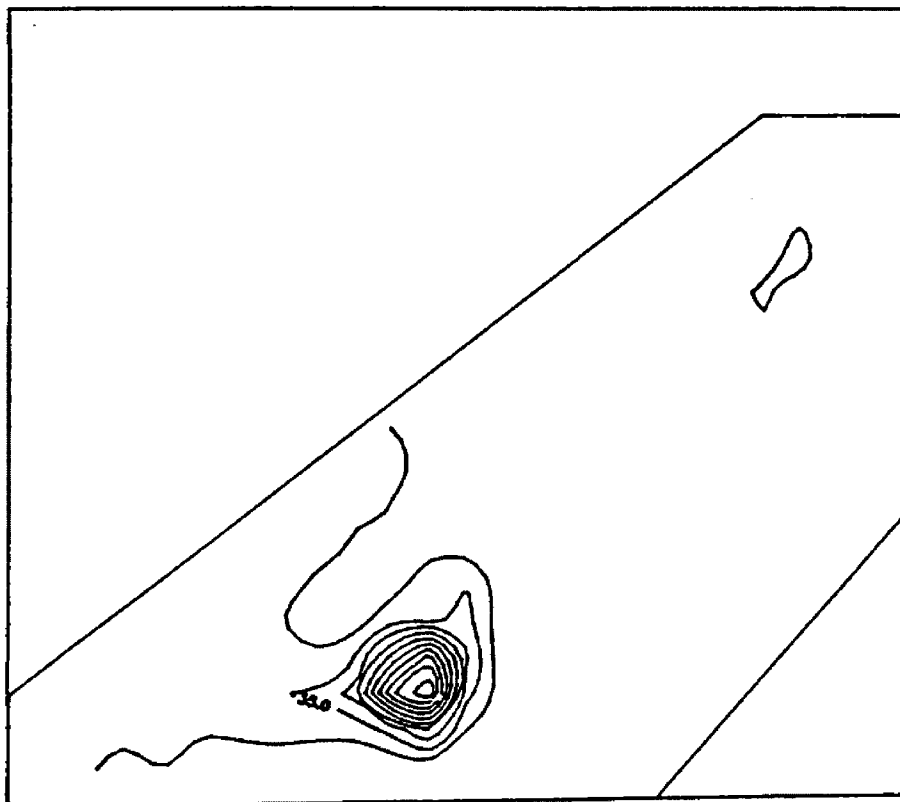
EM 10V DATA PLOT



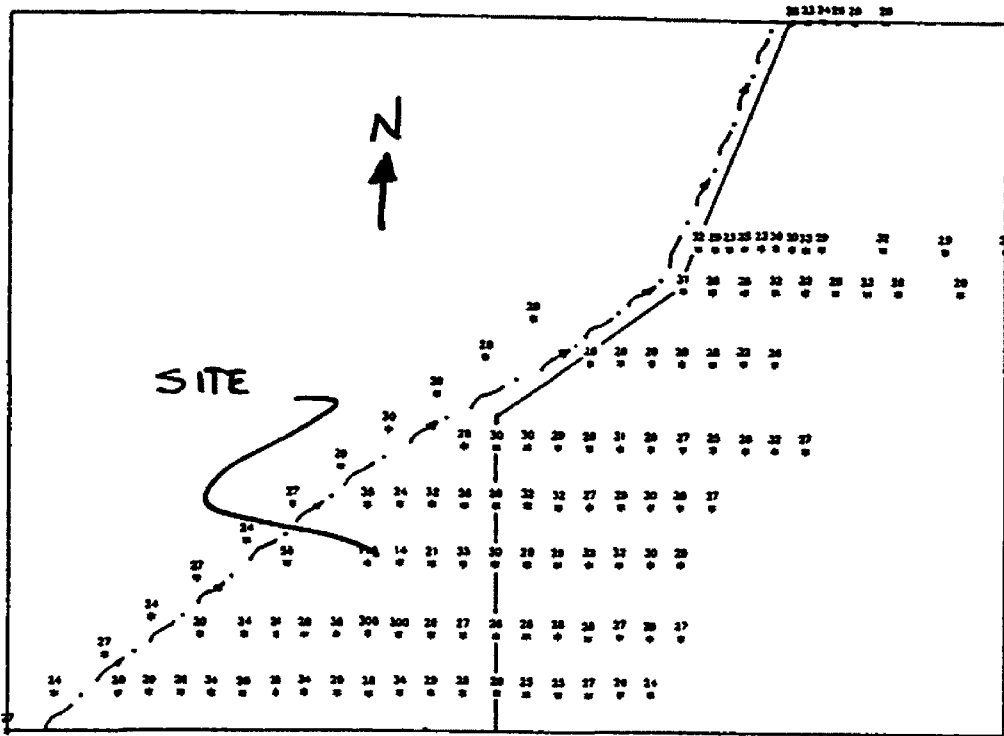
10 METRE EM - HORIZONTAL COIL, $Cl=20$ mmhos/m



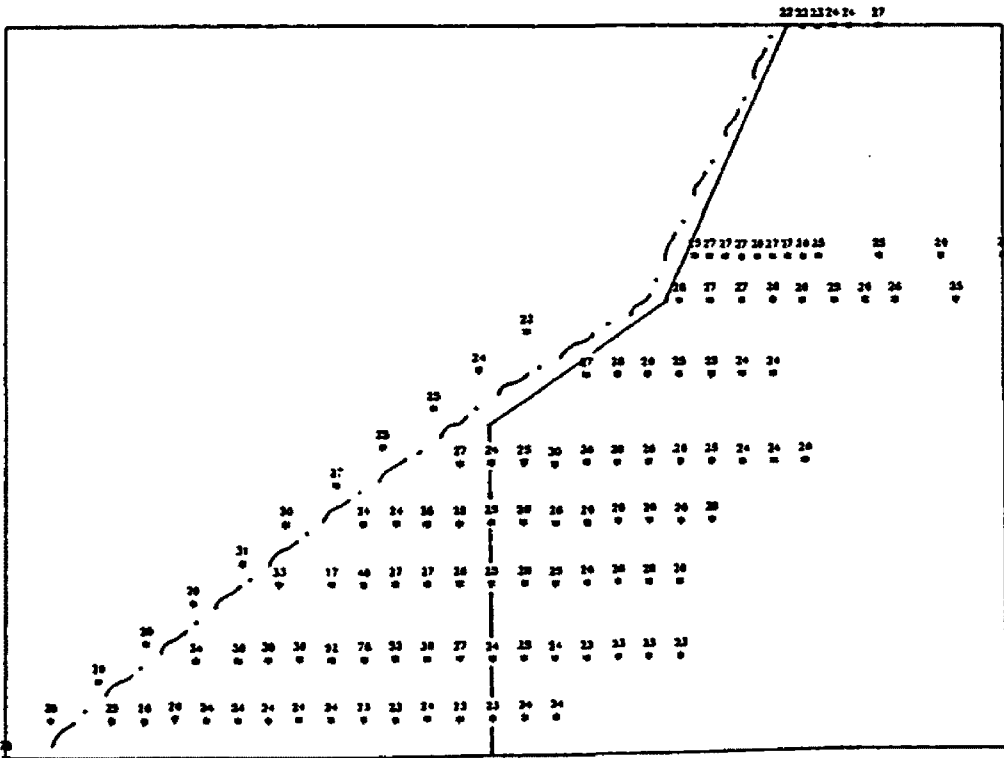
10 METRE EM - VERTICAL COIL, $Cl=10$ mmhos/m



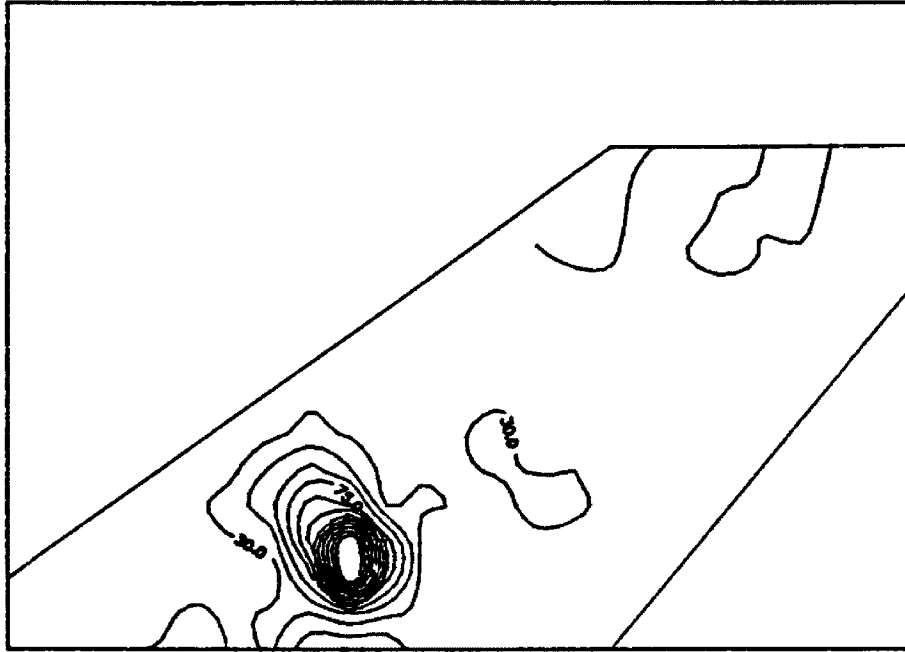
EM 20H DATA PLOT



EM 20V DATA PLOT



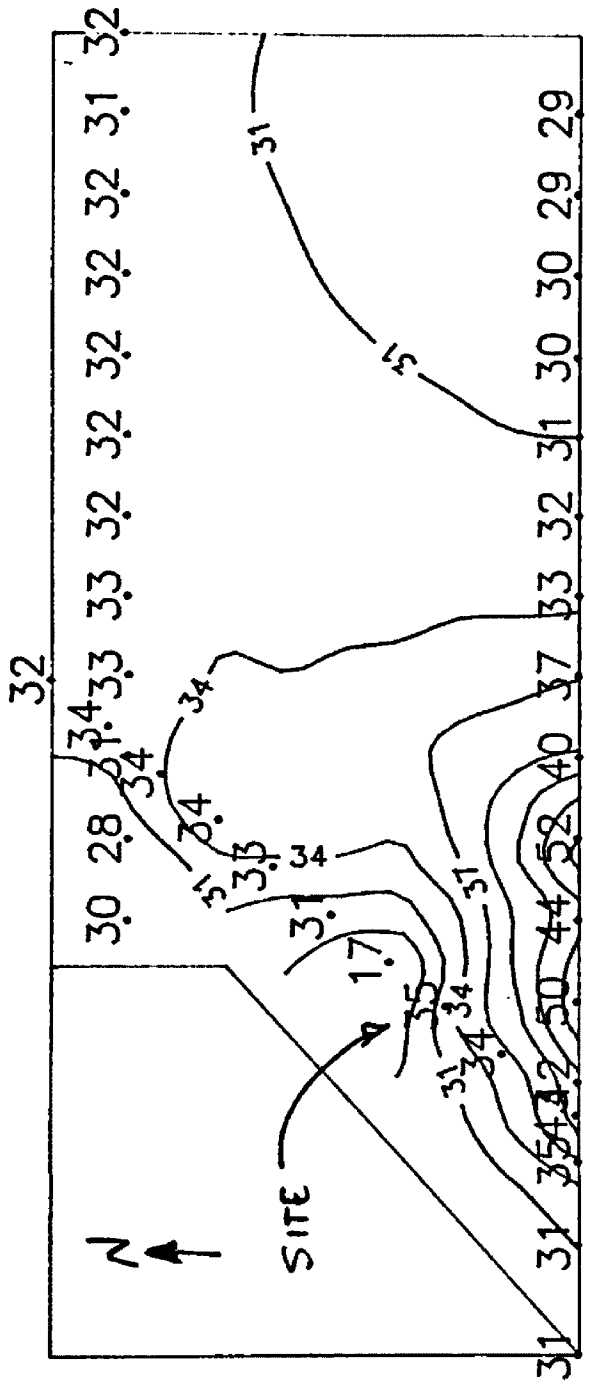
20 METRE EM - HORIZONTAL COIL, CI=15 mmhos/m



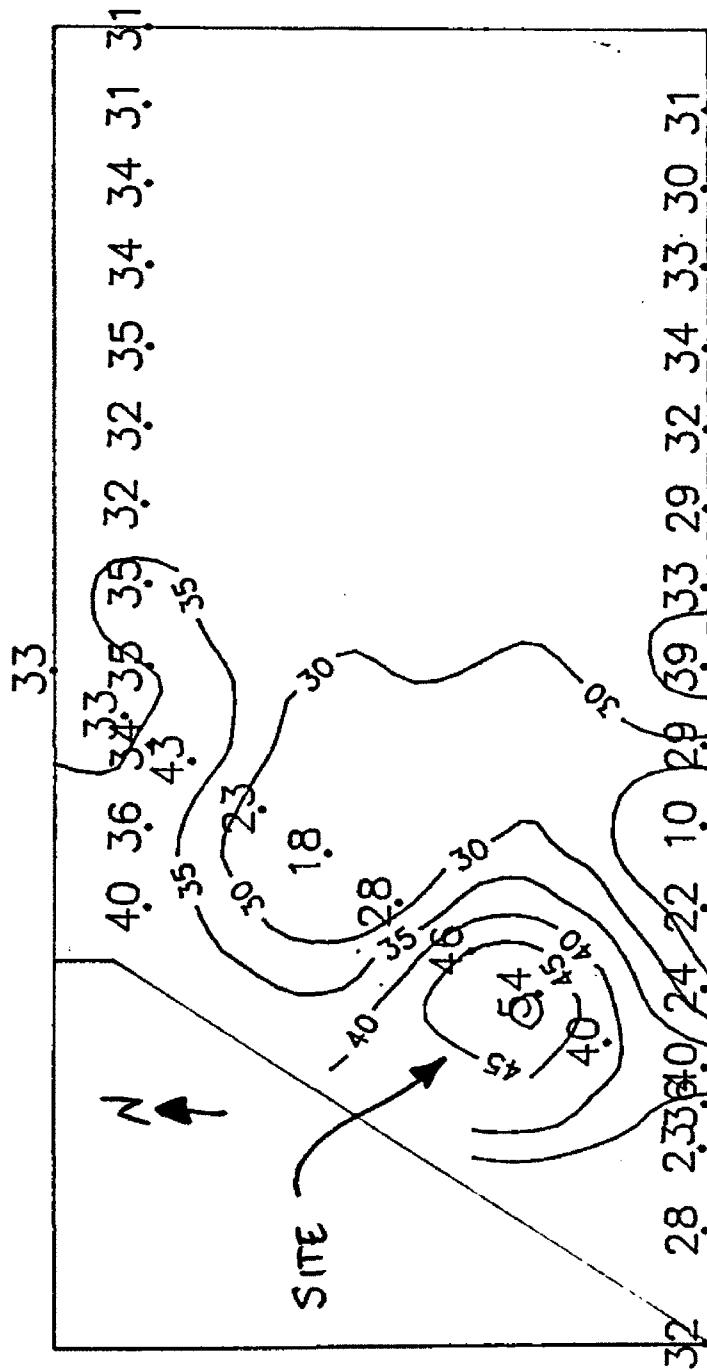
20 METRE EM - VERTICAL COIL, CI=5 mmhos/m



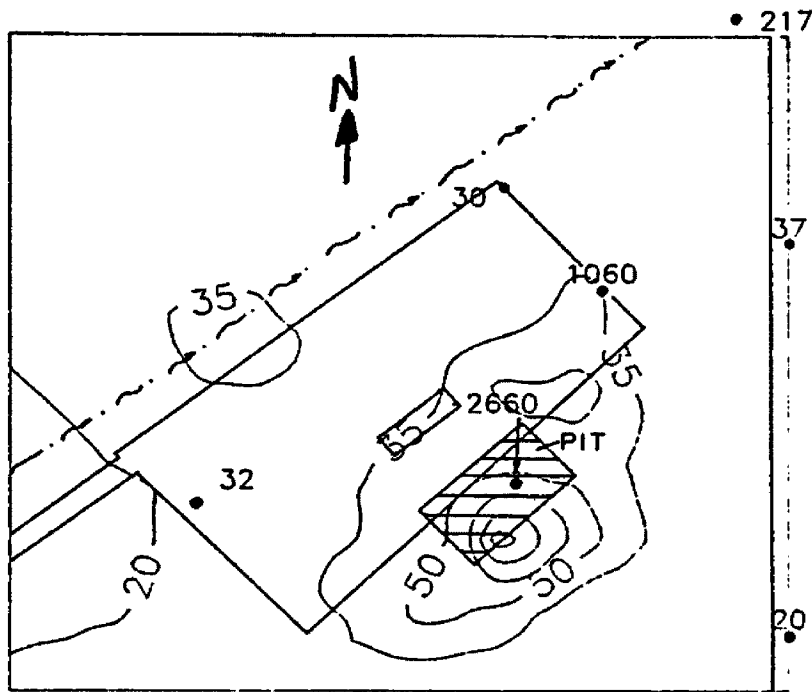
40m EM-VERT. COIL, CI=5 mmohs/m



40m EM-HORZ. COIL, Cl=5 mmohs/m

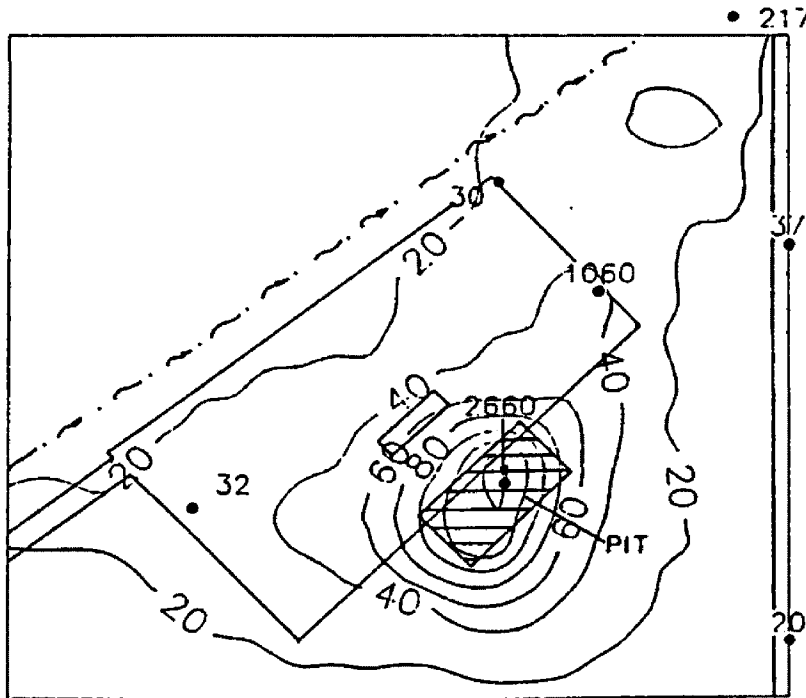


10m EM-HORZ. (mmhos/m) PLOTTED with ppm Cl

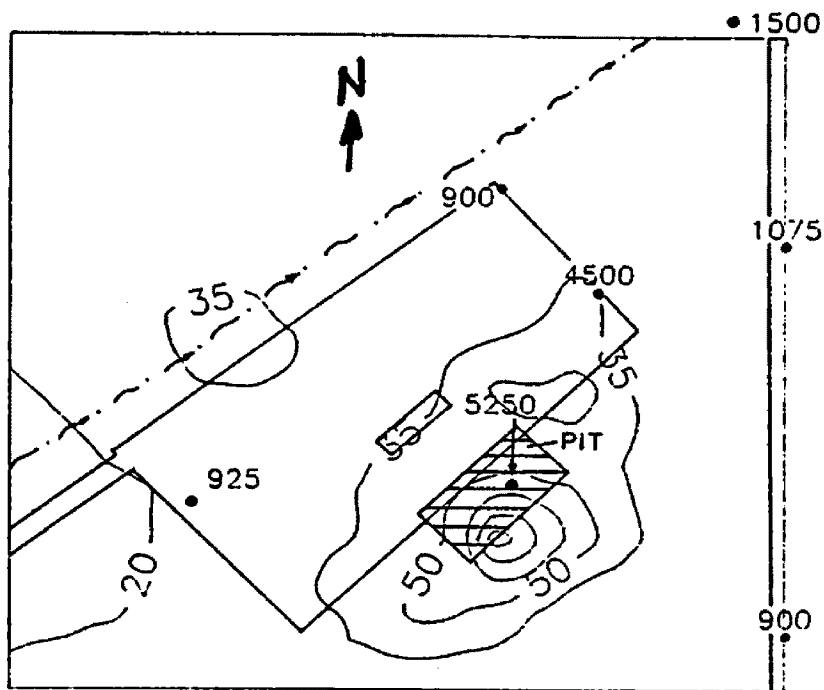


(SMALL #'S ARE Cl)

10m EM-VERT.(mmhos/m) PLOTTED with ppm Cl

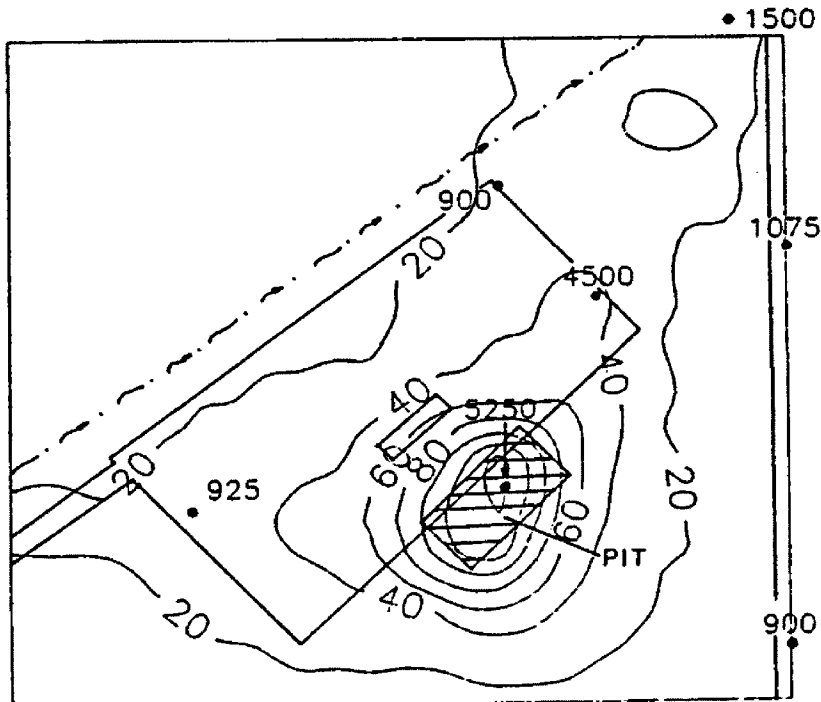


10m EM-HORZ.(mmhos/m) PLOTTED with SC (umhos/cm)



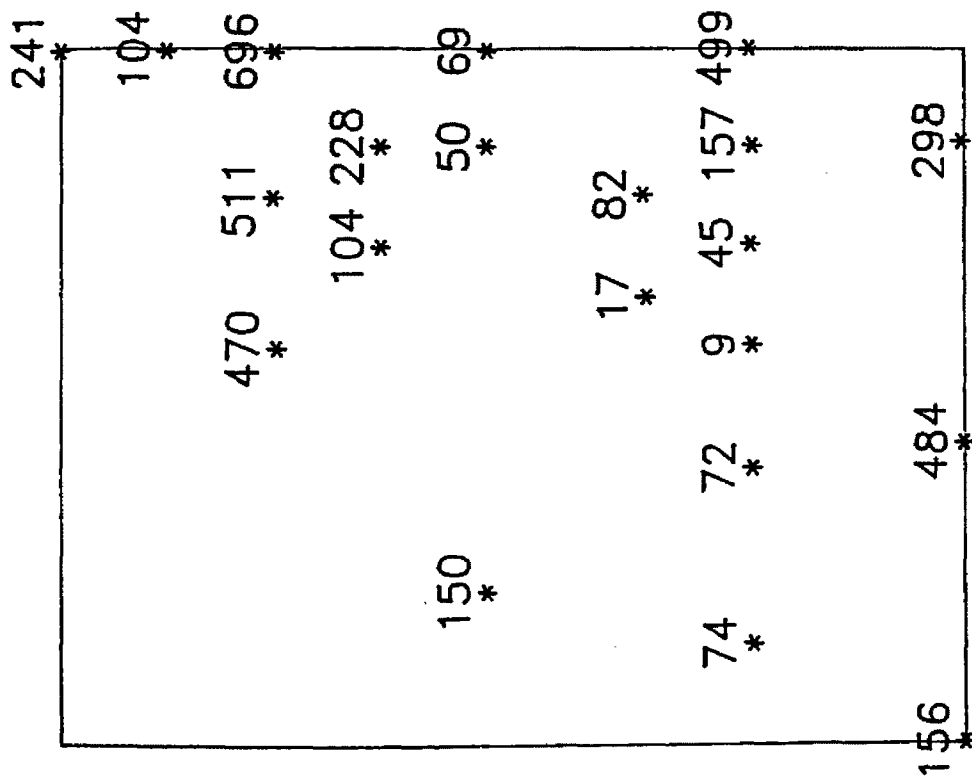
(SMALL #'S ARE SC)

10m EM-VERT.(mmhos/m) PLOTTED with SC (mmhos/cm)

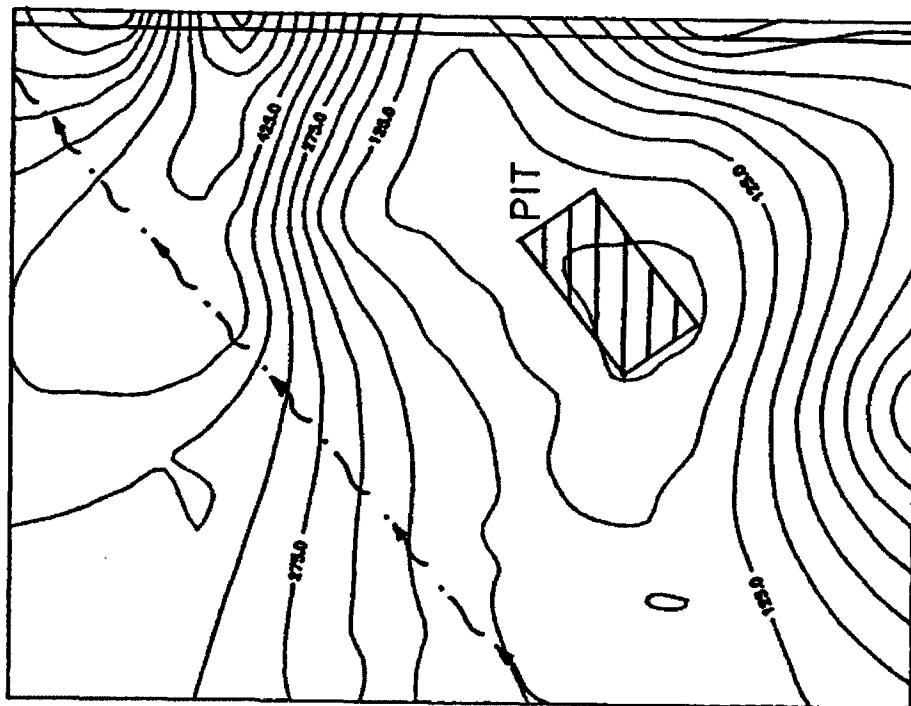


APPENDIX Q
RESISTIVITY DATA/PLOTS

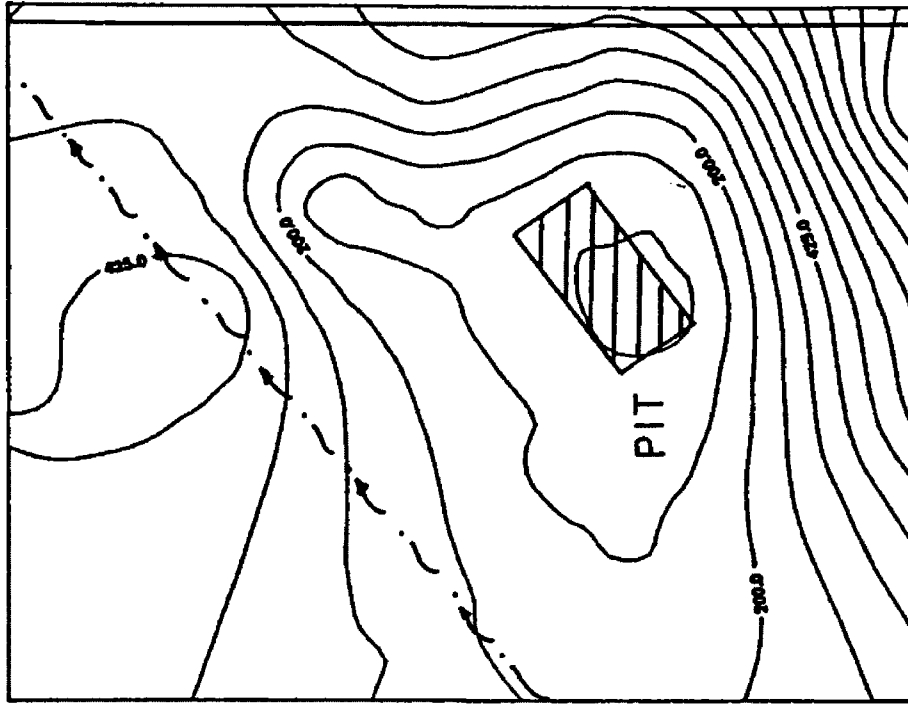
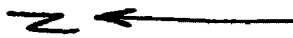
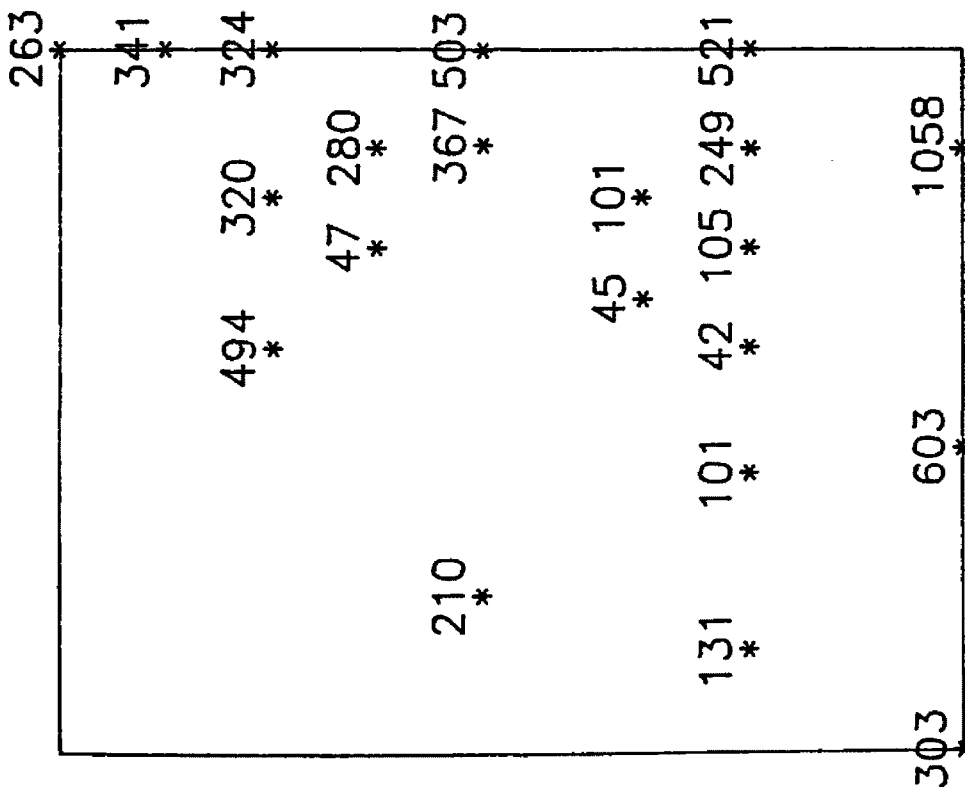
10 ft. σ -SPACING ISORESISTIVITY DATA IN OHM-ft



10 ft. σ -SPACING ISORESISTIVITY, CI=50 OHM-ft



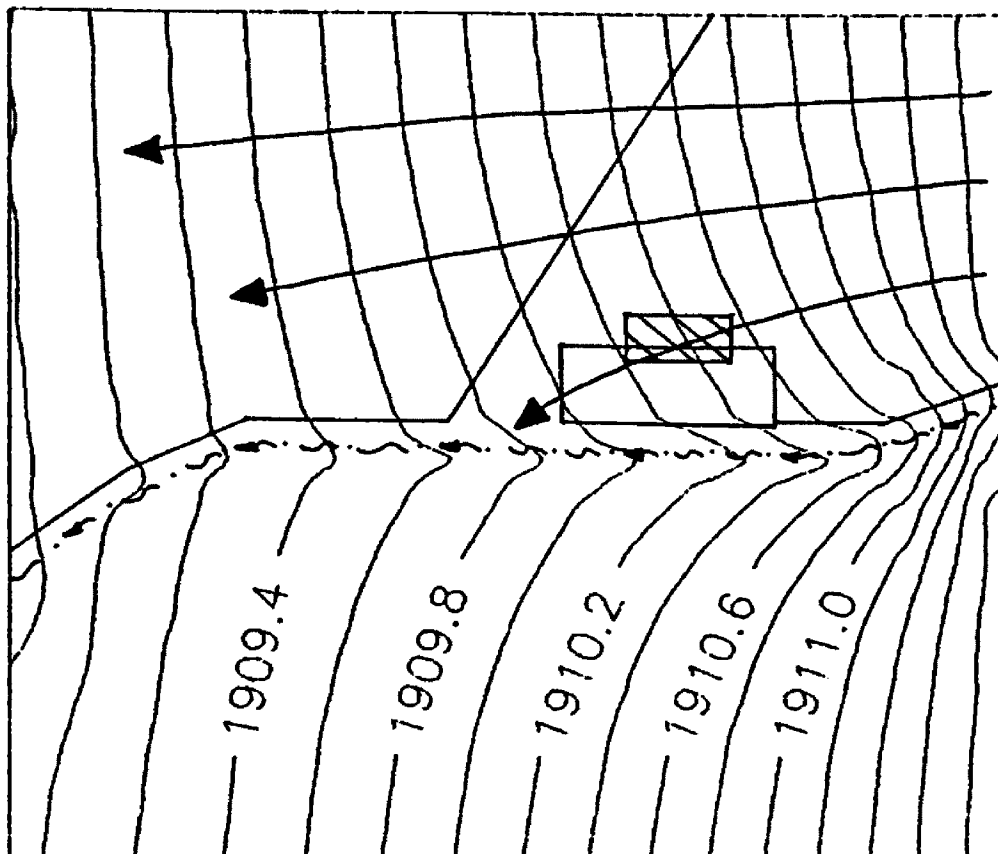
20 ft. α -SPACING ISORESISTIVITY DATA IN OHM-ft



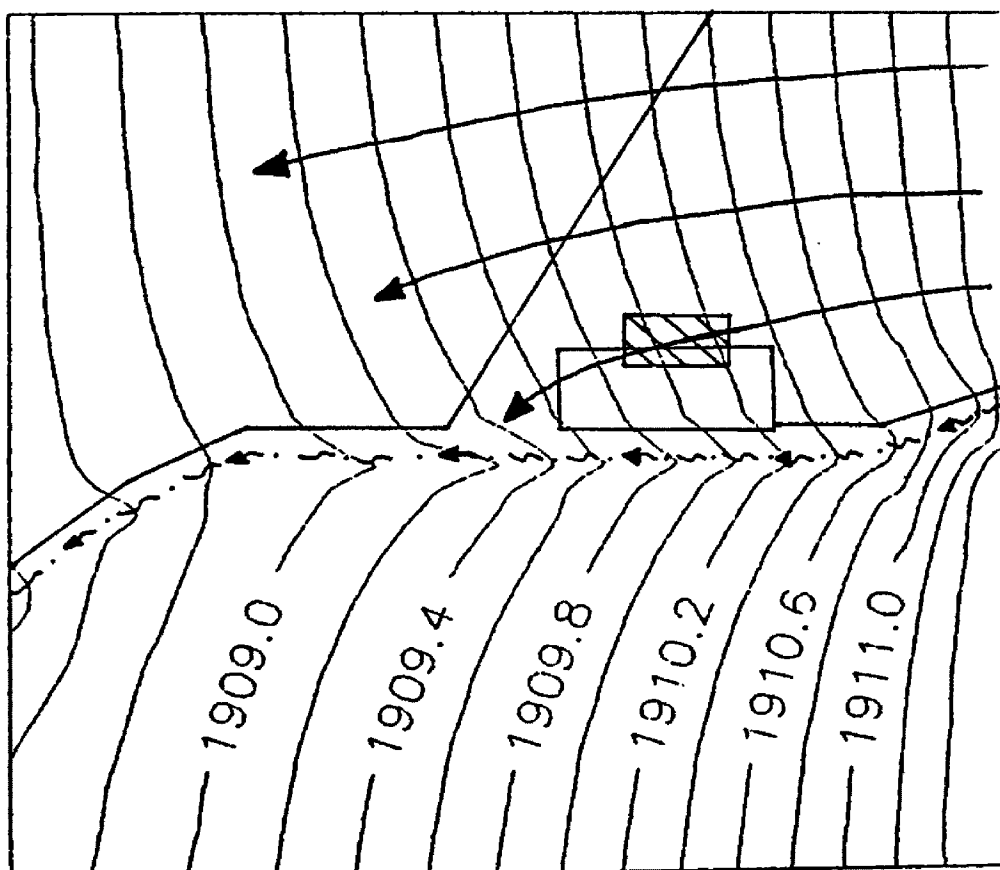
SAME SCALE AS 10 ft. DATA

APPENDIX R
JULY, 87 TO JUNE, 88 SIMULATED FLOW FIELDS

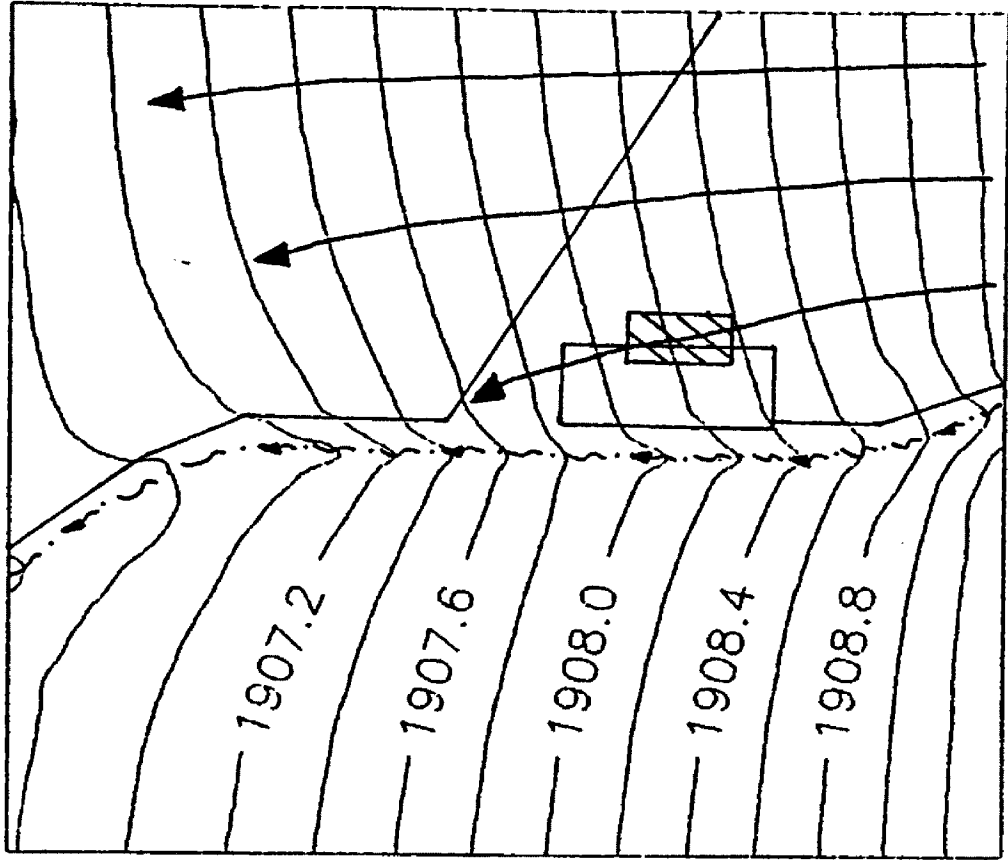
AUGUST



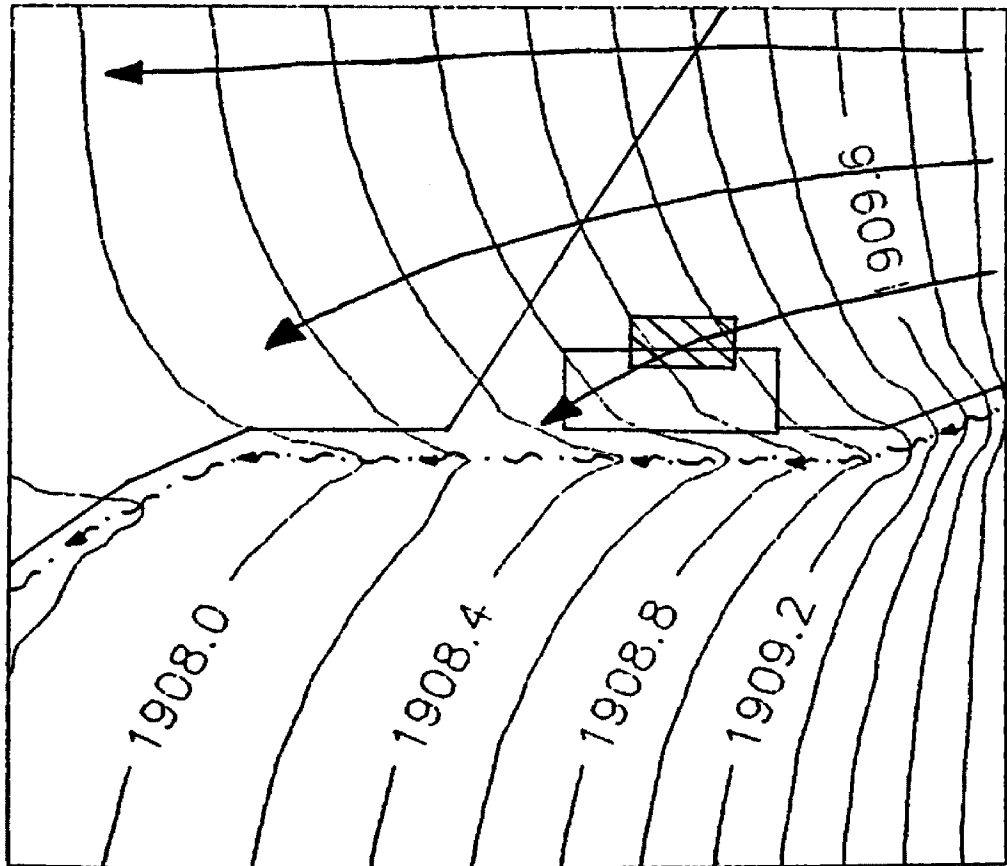
JULY



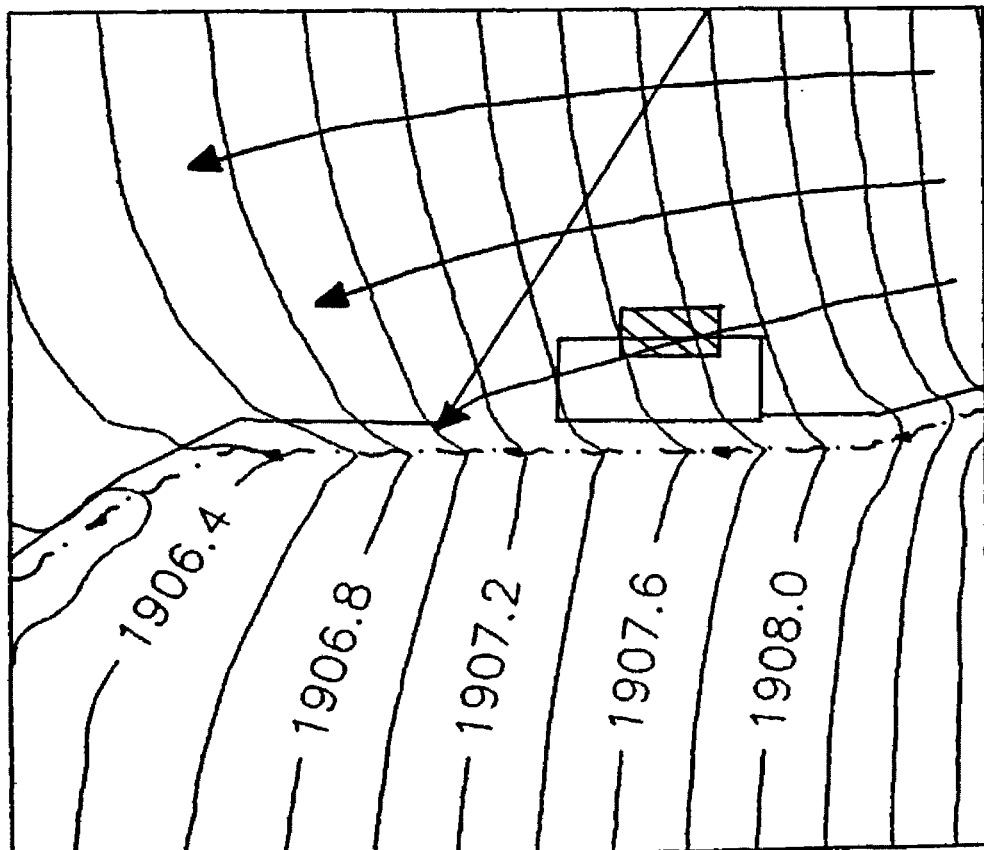
OCTOBER



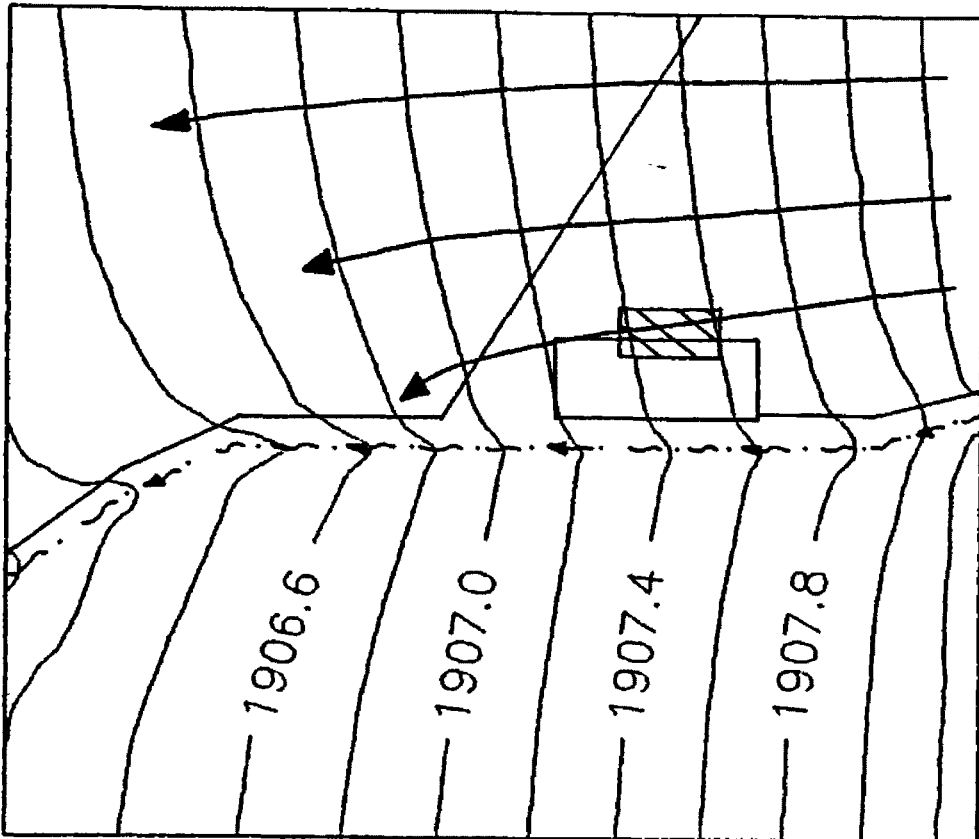
SEPTEMBER ↗



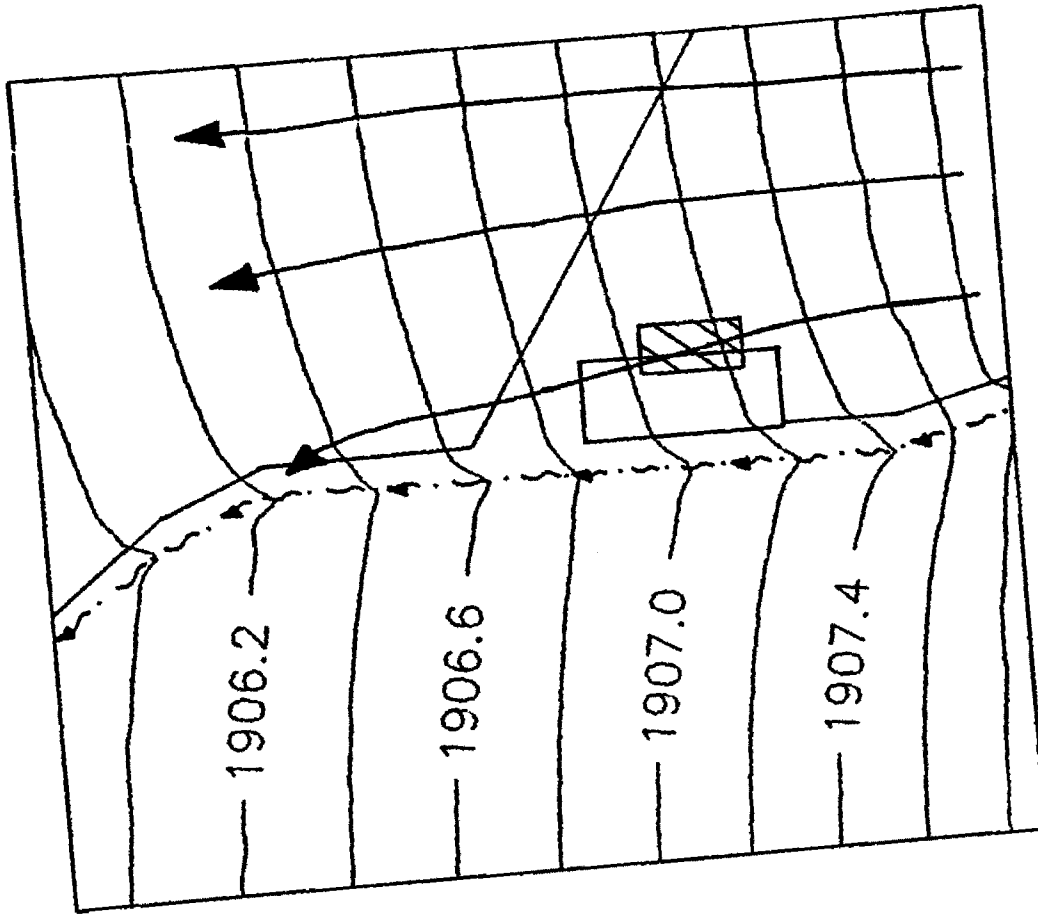
NOVEMBER



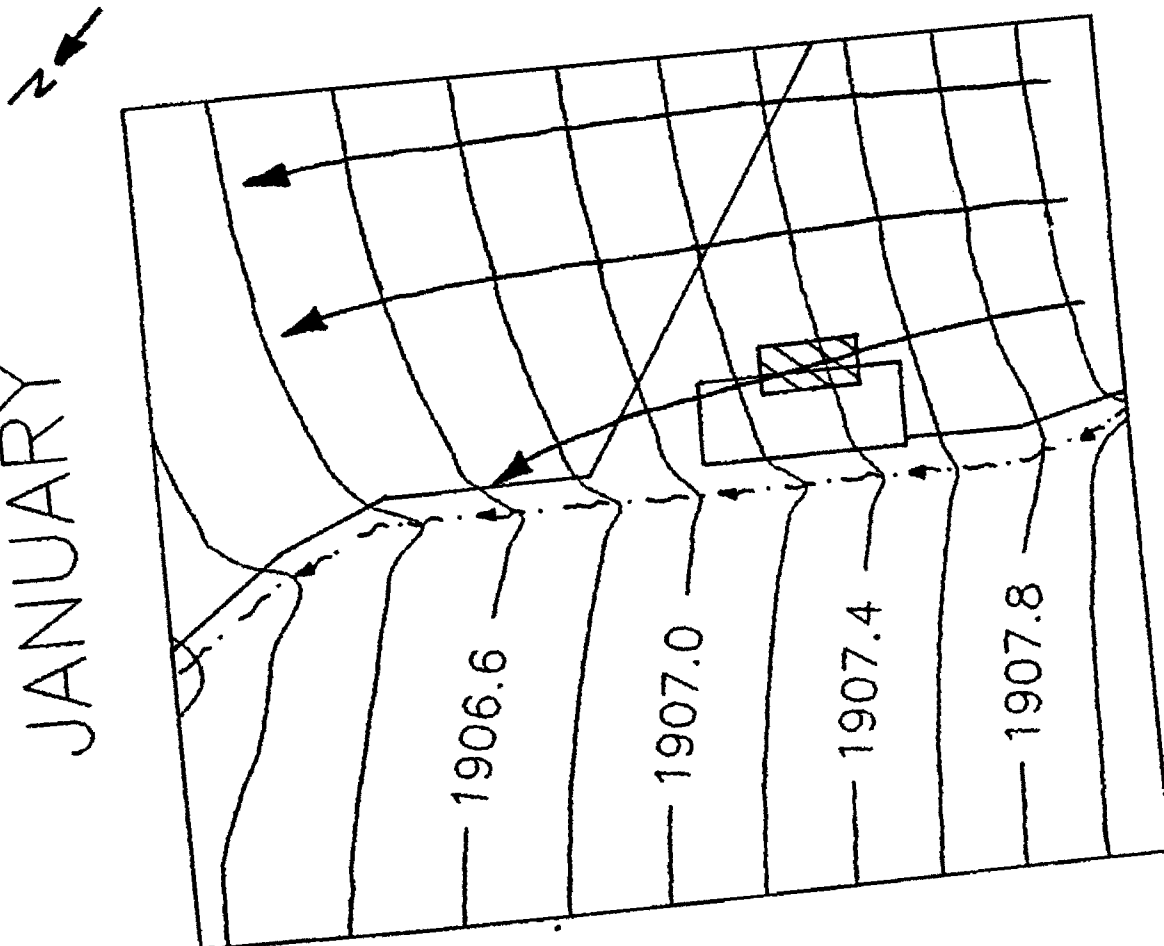
DECEMBER



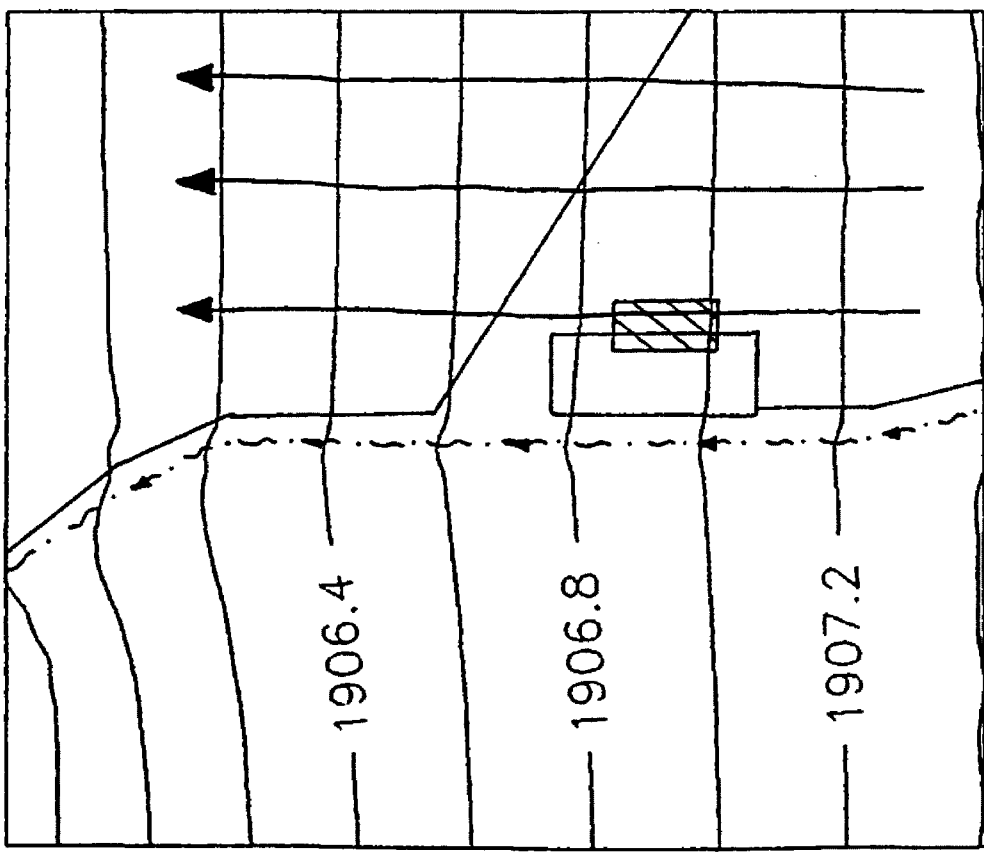
FEBUARY



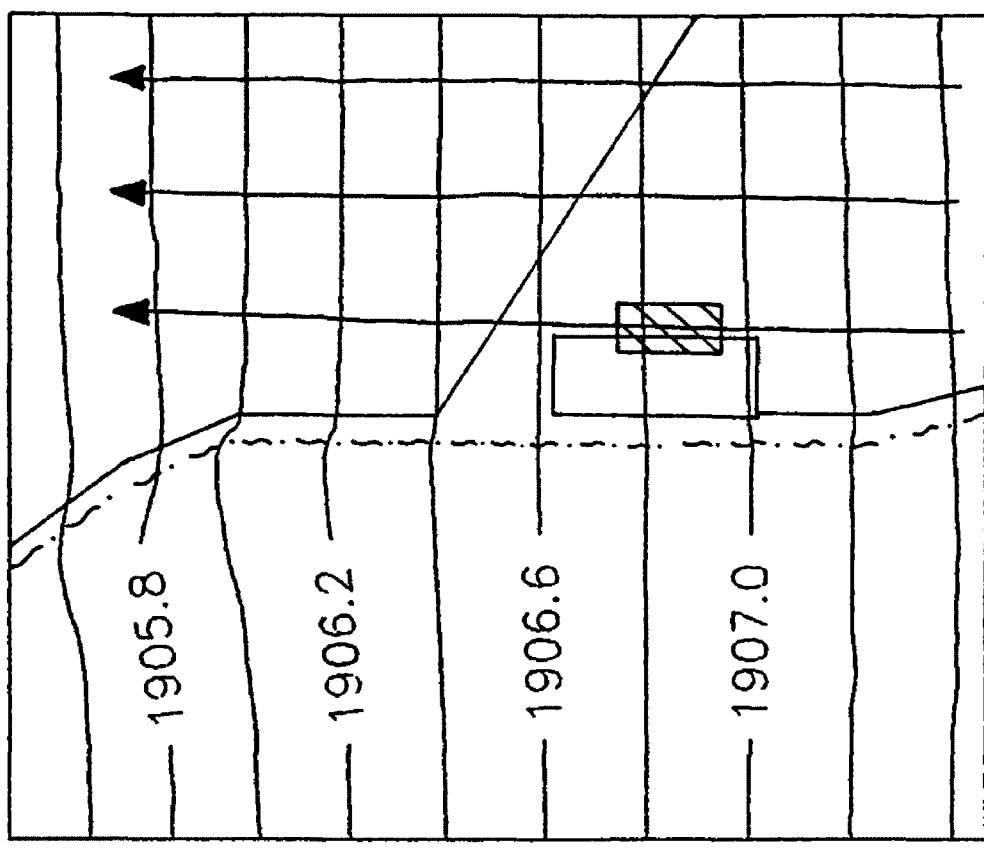
JANUARY



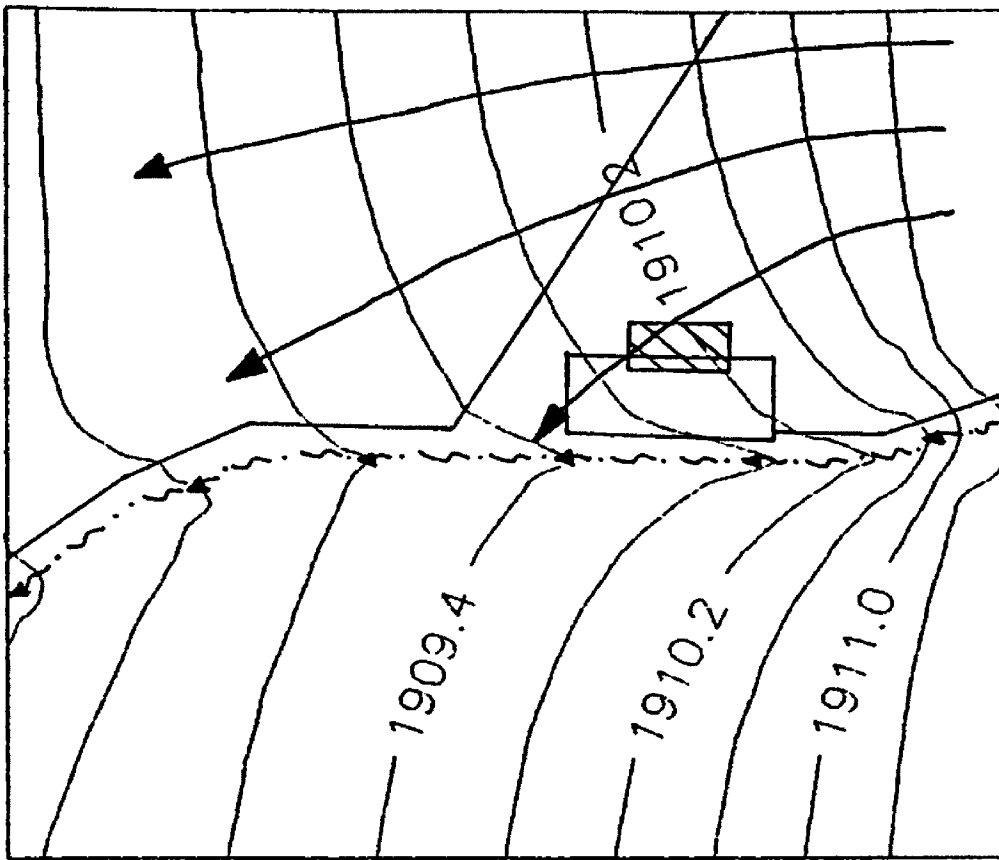
MARCH



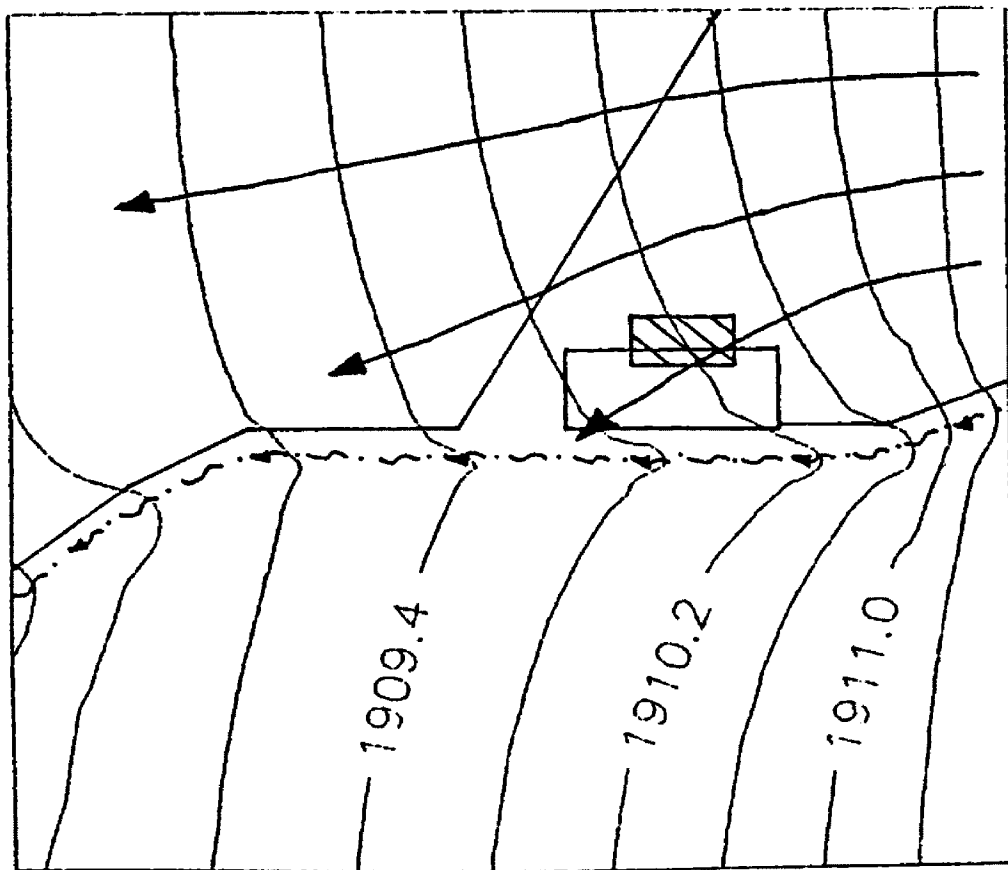
APRIL



JUNE



MAY



APPENDIX S
CONSTANT HEAD DATA FOR MODELING

MONTHLY CONSTANT HEAD INPUT FOR FLOW MODEL

MONTH	i	j	K HEAD	MONTH	i	j	K HEAD	MONTH	i	j	K HEAD
7	1	1	1908.65	8	1	1	1908.93	9	1	1	1907.9
7	2	1	1908.6	8	2	1	1908.88	9	2	1	1907.85
7	3	1	1908.44	8	3	1	1908.75	9	3	1	1907.8
7	4	1	1908.3	8	4	1	1908.65	9	4	1	1907.7
7	5	1	1908.49	8	5	1	1908.75	9	5	1	1907.79
7	6	1	1908.5	8	6	1	1908.76	9	6	1	1907.79
7	7	1	1908.51	8	7	1	1908.76	9	7	1	1907.79
7	8	1	1908.51	8	8	1	1908.76	9	8	1	1907.82
7	9	1	1908.51	8	9	1	1908.76	9	9	1	1907.84
7	10	1	1908.52	8	10	1	1908.77	9	10	1	1907.85
7	11	1	1908.52	8	11	1	1908.77	9	11	1	1907.86
7	12	1	1908.52	8	12	1	1908.77	9	12	1	1907.87
7	13	1	1908.52	8	13	1	1908.77	9	13	1	1907.88
7	14	1	1908.53	8	14	1	1908.78	9	14	1	1907.89
7	15	1	1908.53	8	15	1	1908.78	9	15	1	1907.89
7	16	1	1908.53	8	16	1	1908.78	9	16	1	1907.9
7	17	1	1908.53	8	17	1	1908.78	9	17	1	1907.9
7	18	1	1908.54	8	18	1	1908.79	9	18	1	1907.9
7	1	24	1911.65	8	1	24	1911.95	9	1	24	1910.15
7	2	24	1911.64	8	2	24	1911.94	9	2	24	1910.14
7	3	24	1911.64	8	3	24	1911.94	9	3	24	1910.14
7	4	24	1911.63	8	4	24	1911.93	9	4	24	1910.13
7	5	24	1911.63	8	5	24	1911.93	9	5	24	1910.13
7	6	24	1911.63	8	6	24	1911.93	9	6	24	1910.13
7	7	24	1911.63	8	7	24	1911.93	9	7	24	1910.13
7	8	24	1911.63	8	8	24	1911.93	9	8	24	1910.13
7	9	24	1911.5	8	9	24	1911.8	9	9	24	1910.08
7	10	24	1911.2	8	10	24	1911.4	9	10	24	1910.06
7	11	24	1911	8	11	24	1911.1	9	11	24	1909.15
7	12	24	1911.05	8	12	24	1911.32	9	12	24	1909.95
7	13	24	1911.2	8	13	24	1911.6	9	13	24	1910.08
7	14	24	1911.33	8	14	24	1911.73	9	14	24	1910.08
7	15	24	1911.34	8	15	24	1911.74	9	15	24	1910.09
7	16	24	1911.35	8	16	24	1911.75	9	16	24	1910.1
7	17	24	1911.35	8	17	24	1911.75	9	17	24	1910.11
7	18	24	1911.39	8	18	24	1911.78	9	18	24	1910.13
7	4	1	1908.3	8	4	1	1908.65	9	4	1	1907.7
7	6	2	1908.55	8	6	2	1909	9	6	2	1907.75
7	9	3	1908.75	8	9	3	1909.2	9	9	3	1907.8
7	9	4	1908.82	8	9	4	1909.28	9	9	4	1907.83
7	9	5	1908.87	8	9	5	1909.36	9	9	5	1907.86
7	9	6	1908.9	8	9	6	1909.4	9	9	6	1907.9
7	9	7	1908.93	8	9	7	1909.44	9	9	7	1907.93
7	9	8	1908.95	8	9	8	1909.48	9	9	8	1907.97
7	9	9	1908.99	8	9	9	1909.52	9	9	9	1908
7	9	10	1909.02	8	9	10	1909.58	9	9	10	1908.04
7	9	11	1909.07	8	9	11	1909.63	9	9	11	1908.09
7	9	12	1909.13	8	9	12	1909.69	9	9	12	1908.13
7	9	13	1909.26	8	9	13	1909.75	9	9	13	1908.19
7	9	14	1909.39	8	9	14	1909.79	9	9	14	1908.21
7	9	15	1909.5	8	9	15	1909.82	9	9	15	1908.24
7	9	16	1909.6	8	9	16	1909.86	9	9	16	1908.27
7	9	17	1909.7	8	9	17	1909.9	9	9	17	1908.34
7	9	18	1909.9	8	9	18	1909.95	9	9	18	1908.4
7	9	19	1909.9	8	9	19	1909.99	9	9	19	1908.45
7	9	20	1910	8	9	20	1910.02	9	9	20	1908.5
7	9	21	1910.1	8	9	21	1910.06	9	9	21	1908.55
7	9	22	1910.22	8	9	22	1910.11	9	9	22	1908.63
7	9	23	1910.4	8	9	23	1910.16	9	9	23	1908.8
7	11	24	1911	8	11	24	1911.1	9	11	24	1909.15

MONTH	i	j	K HEAD	MONTH	i	j	K HEAD	MONTH	i	j	K HEAD
10	1	1	1906.78	11	1	1	1906.3	12	1	1	1906.23
10	2	1	1906.75	11	2	1	1906.29	12	2	1	1906.22
10	3	1	1906.73	11	3	1	1906.28	12	3	1	1906.08
10	4	1	1906.53	11	4	1	1905.88	12	4	1	1905.94
10	5	1	1906.73	11	5	1	1906.29	12	5	1	1906.08
10	6	1	1906.74	11	6	1	1906.29	12	6	1	1906.2
10	7	1	1906.75	11	7	1	1906.29	12	7	1	1906.2
10	8	1	1906.75	11	8	1	1906.29	12	8	1	1906.2
10	9	1	1906.75	11	9	1	1906.29	12	9	1	1906.21
10	10	1	1906.75	11	10	1	1906.3	12	10	1	1906.21
10	11	1	1906.76	11	11	1	1906.31	12	11	1	1906.22
10	12	1	1906.76	11	12	1	1906.33	12	12	1	1906.22
10	13	1	1906.76	11	13	1	1906.35	12	13	1	1906.23
10	14	1	1906.77	11	14	1	1906.37	12	14	1	1906.25
10	15	1	1906.77	11	15	1	1906.39	12	15	1	1906.27
10	16	1	1906.78	11	16	1	1906.41	12	16	1	1906.29
10	17	1	1906.79	11	17	1	1906.43	12	17	1	1906.31
10	18	1	1906.8	11	18	1	1906.45	12	18	1	1906.33
10	1	24	1909.37	11	1	24	1908.7	12	1	24	1908.3
10	2	24	1909.36	11	2	24	1908.69	12	2	24	1908.29
10	3	24	1909.36	11	3	24	1908.69	12	3	24	1908.29
10	4	24	1909.35	11	4	24	1908.68	12	4	24	1908.29
10	5	24	1909.35	11	5	24	1908.68	12	5	24	1908.29
10	6	24	1909.35	11	6	24	1908.68	12	6	24	1908.29
10	7	24	1909.35	11	7	24	1908.68	12	7	24	1908.28
10	8	24	1909.35	11	8	24	1908.68	12	8	24	1908.28
10	9	24	1909.3	11	9	24	1908.63	12	9	24	1908.26
10	10	24	1909.2	11	10	24	1908.53	12	10	24	1908.24
10	11	24	1908.66	11	11	24	1908.19	12	11	24	1907.84
10	12	24	1909.23	11	12	24	1908.7	12	12	24	1908.26
10	13	24	1909.25	11	13	24	1908.72	12	13	24	1908.27
10	14	24	1909.25	11	14	24	1908.74	12	14	24	1908.28
10	15	24	1909.26	11	15	24	1909.75	12	15	24	1908.3
10	16	24	1909.27	11	16	24	1908.78	12	16	24	1908.32
10	17	24	1909.28	11	17	24	1908.8	12	17	24	1908.34
10	18	24	1909.3	11	18	24	1908.82	12	18	24	1908.36
10	4	1	1906.53	11	4	1	1905.88	12	4	1	1905.94
10	6	2	1906.61	11	6	2	1906.06	12	6	2	1906.14
10	9	3	1906.69	11	9	3	1906.22	12	9	3	1906.24
10	9	4	1906.77	11	9	4	1906.3	12	9	4	1906.31
10	9	5	1906.85	11	9	5	1906.38	12	9	5	1906.4
10	9	6	1906.97	11	9	6	1906.5	12	9	6	1906.45
10	9	7	1907.09	11	9	7	1906.62	12	9	7	1906.56
10	9	8	1907.21	11	9	8	1906.74	12	9	8	1906.64
10	9	9	1907.33	11	9	9	1906.86	12	9	9	1906.78
10	9	10	1907.45	11	9	10	1906.98	12	9	10	1906.97
10	9	11	1907.57	11	9	11	1907.1	12	9	11	1906.95
10	9	12	1907.69	11	9	12	1907.19	12	9	12	1907.03
10	9	13	1907.78	11	9	13	1907.26	12	9	13	1907.1
10	9	14	1907.84	11	9	14	1907.34	12	9	14	1907.17
10	9	15	1907.87	11	9	15	1907.4	12	9	15	1907.24
10	9	16	1907.9	11	9	16	1907.43	12	9	16	1907.31
10	9	17	1907.97	11	9	17	1907.5	12	9	17	1907.37
10	9	18	1908.05	11	9	18	1907.6	12	9	18	1907.44
10	9	19	1908.13	11	9	19	1907.66	12	9	19	1907.5
10	9	20	1908.22	11	9	20	1907.73	12	9	20	1907.55
10	9	21	1908.32	11	9	21	1907.75	12	9	21	1907.61
10	9	22	1908.43	11	9	22	1907.96	12	9	22	1907.66
10	9	23	1908.55	11	9	23	1907.98	12	9	23	1907.75
10	11	24	1908.66	11	11	24	1908.19	12	11	24	1907.84

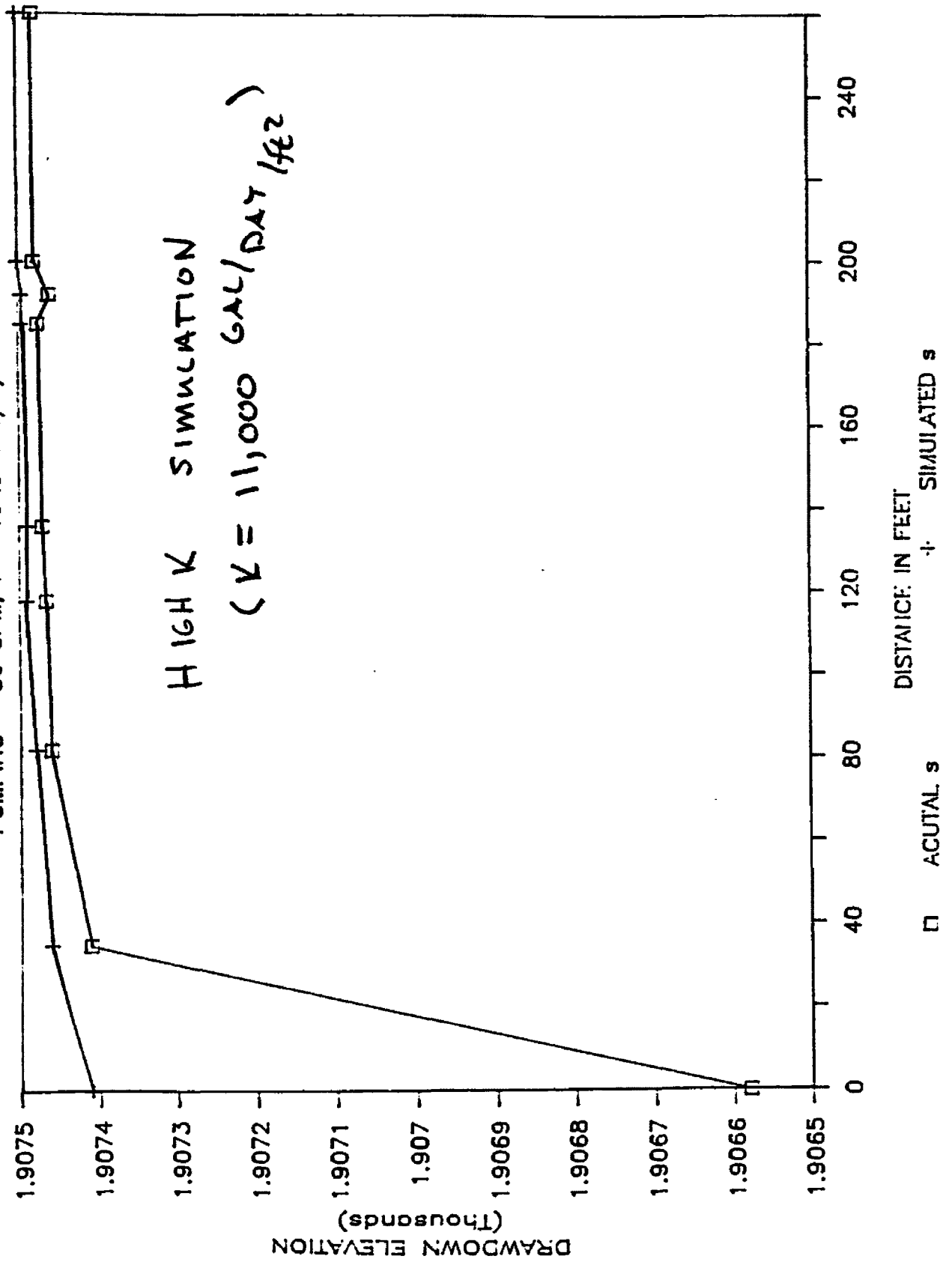
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1	3	1	1906.03	2	3	1	1905.88	3	3	1	1905.69
1	4	1	1905.9	2	4	1	1905.76	3	4	1	1905.82
1	5	1	1906.03	2	5	1	1905.92	3	5	1	1905.79
1	6	1	1906.15	2	6	1	1905.93	3	6	1	1905.79
1	7	1	1906.15	2	7	1	1905.94	3	7	1	1905.79
1	8	1	1906.15	2	8	1	1905.94	3	8	1	1905.9
1	9	1	1906.16	2	9	1	1905.95	3	9	1	1905.8
1	10	1	1906.16	2	10	1	1905.95	3	10	1	1905.81
1	11	1	1906.17	2	11	1	1905.96	3	11	1	1905.81
1	12	1	1906.18	2	12	1	1905.96	3	12	1	1905.81
1	13	1	1906.19	2	13	1	1905.97	3	13	1	1905.92
1	14	1	1906.2	2	14	1	1905.98	3	14	1	1905.82
1	15	1	1906.22	2	15	1	1905.99	3	15	1	1905.83
1	16	1	1906.24	2	16	1	1906	3	16	1	1905.83
1	17	1	1906.26	2	17	1	1906.01	3	17	1	1905.84
1	18	1	1906.28	2	18	1	1906.02	3	18	1	1905.84
1	1	24	1908.15	2	1	24	1907.88	3	1	24	1907.41
1	2	24	1908.14	2	2	24	1907.87	3	2	24	1907.41
1	3	24	1908.14	2	3	24	1907.86	3	3	24	1907.41
1	4	24	1908.13	2	4	24	1907.85	3	4	24	1907.41
1	5	24	1908.13	2	5	24	1907.84	3	5	24	1907.41
1	6	24	1908.13	2	6	24	1907.83	3	6	24	1907.41
1	7	24	1908.13	2	7	24	1907.82	3	7	24	1907.41
1	8	24	1908.12	2	8	24	1907.81	3	8	24	1907.4
1	9	24	1908.1	2	9	24	1907.8	3	9	24	1907.4
1	10	24	1908.05	2	10	24	1907.8	3	10	24	1907.4
1	11	24	1907.91	2	11	24	1907.7	3	11	24	1907.44
1	12	24	1908.07	2	12	24	1907.87	3	12	24	1907.4
1	13	24	1908.08	2	13	24	1907.88	3	13	24	1907.4
1	14	24	1908.1	2	14	24	1907.89	3	14	24	1907.4
1	15	24	1908.12	2	15	24	1907.9	3	15	24	1907.41
1	16	24	1908.14	2	16	24	1907.91	3	16	24	1907.41
1	17	24	1908.16	2	17	24	1907.92	3	17	24	1907.41
1	18	24	1908.18	2	18	24	1907.93	3	18	24	1907.42
1	4	1	1905.9	2	4	1	1905.76	3	4	1	1905.82
1	6	2	1906.13	2	6	2	1905.98	3	6	2	1906.03
1	9	3	1906.27	2	9	3	1906.13	3	9	3	1906.24
1	9	4	1906.39	2	9	4	1906.23	3	9	4	1906.33
1	9	5	1906.43	2	9	5	1906.35	3	9	5	1906.39
1	9	6	1906.51	2	9	6	1906.4	3	9	6	1906.45
1	9	7	1906.6	2	9	7	1906.45	3	9	7	1906.51
1	9	8	1906.65	2	9	8	1906.5	3	9	8	1906.56
1	9	9	1906.7	2	9	9	1906.55	3	9	9	1906.61
1	9	10	1906.76	2	9	10	1906.6	3	9	10	1906.66
1	9	11	1906.84	2	9	11	1906.68	3	9	11	1906.69
1	9	12	1906.92	2	9	12	1906.75	3	9	12	1906.75
1	9	13	1906.98	2	9	13	1906.8	3	9	13	1906.79
1	9	14	1907.04	2	9	14	1906.84	3	9	14	1906.84
1	9	15	1907.09	2	9	15	1906.89	3	9	15	1906.88
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1	9	17	1907.19	2	9	17	1907	3	9	17	1906.96
1	9	18	1907.29	2	9	18	1907.03	3	9	18	1907
1	9	19	1907.35	2	9	19	1907.06	3	9	19	1907.04
1	9	20	1907.41	2	9	20	1907.13	3	9	20	1907.08
1	9	21	1907.51	2	9	21	1907.2	3	9	21	1907.12
1	9	22	1907.59	2	9	22	1907.25	3	9	22	1907.17
1	9	23	1907.71	2	9	23	1907.34	3	9	23	1907.26
1	11	24	1907.91	2	11	24	1907.7	3	11	24	1907.42

MONTH	i	j	K HEAD	MONTH	i	j	K HEAD	MONTH	i	j	K HEAD
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4	2	1	1906	5	2	1	1908.45	6	2	1	1908.45
4	3	1	1906	5	3	1	1908.4	6	3	1	1908.25
4	4	1	1906	5	4	1	1907.98	6	4	1	1907.89
4	5	1	1906	5	5	1	1908.52	6	5	1	1908.45
4	6	1	1906	5	6	1	1908.53	6	6	1	1908.46
4	7	1	1906	5	7	1	1908.54	6	7	1	1908.47
4	8	1	1906	5	8	1	1908.55	6	8	1	1908.48
4	9	1	1906	5	9	1	1908.57	6	9	1	1908.5
4	10	1	1906	5	10	1	1908.6	6	10	1	1908.53
4	11	1	1906	5	11	1	1908.62	6	11	1	1908.55
4	12	1	1906	5	12	1	1908.62	6	12	1	1908.55
4	13	1	1906	5	13	1	1908.62	6	13	1	1908.55
4	14	1	1906	5	14	1	1908.61	6	14	1	1908.54
4	15	1	1906	5	15	1	1908.59	6	15	1	1908.52
4	16	1	1906	5	16	1	1908.57	6	16	1	1908.5
4	17	1	1906	5	17	1	1908.57	6	17	1	1908.5
4	18	1	1906	5	18	1	1908.55	6	18	1	1908.48
4	1	24	1908	5	1	24	1911.82	6	1	24	1911.97
4	2	24	1908	5	2	24	1911.79	6	2	24	1911.94
4	3	24	1908	5	3	24	1911.79	6	3	24	1911.94
4	4	24	1908	5	4	24	1911.77	6	4	24	1911.92
4	5	24	1908	5	5	24	1911.77	6	5	24	1911.92
4	6	24	1908	5	6	24	1911.76	6	6	24	1911.91
4	7	24	1908	5	7	24	1911.76	6	7	24	1911.91
4	8	24	1908	5	8	24	1911.75	6	8	24	1911.91
4	9	24	1908	5	9	24	1911.7	6	9	24	1911.85
4	10	24	1908	5	10	24	1911.6	6	10	24	1911.75
4	11	24	1909	5	11	24	1910.85	6	11	24	1909.54
4	12	24	1908	5	12	24	1911.66	6	12	24	1911.91
4	13	24	1908	5	13	24	1911.68	6	13	24	1911.93
4	14	24	1908	5	14	24	1911.7	6	14	24	1911.95
4	15	24	1908	5	15	24	1911.7	6	15	24	1911.95
4	16	24	1908	5	16	24	1911.66	6	16	24	1911.91
4	17	24	1908	5	17	24	1911.65	6	17	24	1911.9
4	18	24	1908	5	18	24	1911.65	6	18	24	1911.9
4	4	1	1906	5	4	1	1907.98	6	4	1	1907.98
4	6	2	1906.24	5	6	2	1908.4	6	6	2	1908.28
4	9	3	1906.4	5	9	3	1908.72	6	9	3	1908.58
4	9	4	1906.48	5	9	4	1908.98	6	9	4	1908.78
4	9	5	1906.57	5	9	5	1909.12	6	9	5	1908.92
4	9	6	1906.68	5	9	6	1909.16	6	9	6	1908.96
4	9	7	1906.8	5	9	7	1909.19	6	9	7	1908.99
4	9	8	1906.92	5	9	8	1909.22	6	9	8	1909.02
4	9	9	1907.04	5	9	9	1909.31	6	9	9	1909.06
4	9	10	1906.16	5	9	10	1909.36	6	9	10	1909.11
4	9	11	1907.28	5	9	11	1909.4	6	9	11	1909.15
4	9	12	1907.4	5	9	12	1909.44	6	9	12	1909.19
4	9	13	1907.47	5	9	13	1909.47	6	9	13	1909.22
4	9	14	1907.56	5	9	14	1909.5	6	9	14	1909.25
4	9	15	1907.58	5	9	15	1909.53	6	9	15	1909.26
4	9	16	1907.61	5	9	16	1909.58	6	9	16	1909.33
4	9	17	1907.68	5	9	17	1909.63	6	9	17	1909.38
4	9	18	1907.78	5	9	18	1909.66	6	9	18	1909.41
4	9	19	1907.84	5	9	19	1909.7	6	9	19	1909.45
4	9	20	1907.93	5	9	20	1909.73	6	9	20	1909.48
4	9	21	1908.03	5	9	21	1909.76	6	9	21	1909.51
4	9	22	1908.14	5	9	22	1909.95	6	9	22	1909.55
4	9	23	1908.28	5	9	23	1910.1	6	9	23	1909.58
6	11	24	1908.37	5	11	24	1910.85	6	11	24	1909.64

APPENDIX T
FLOW MODELING STATISTICS

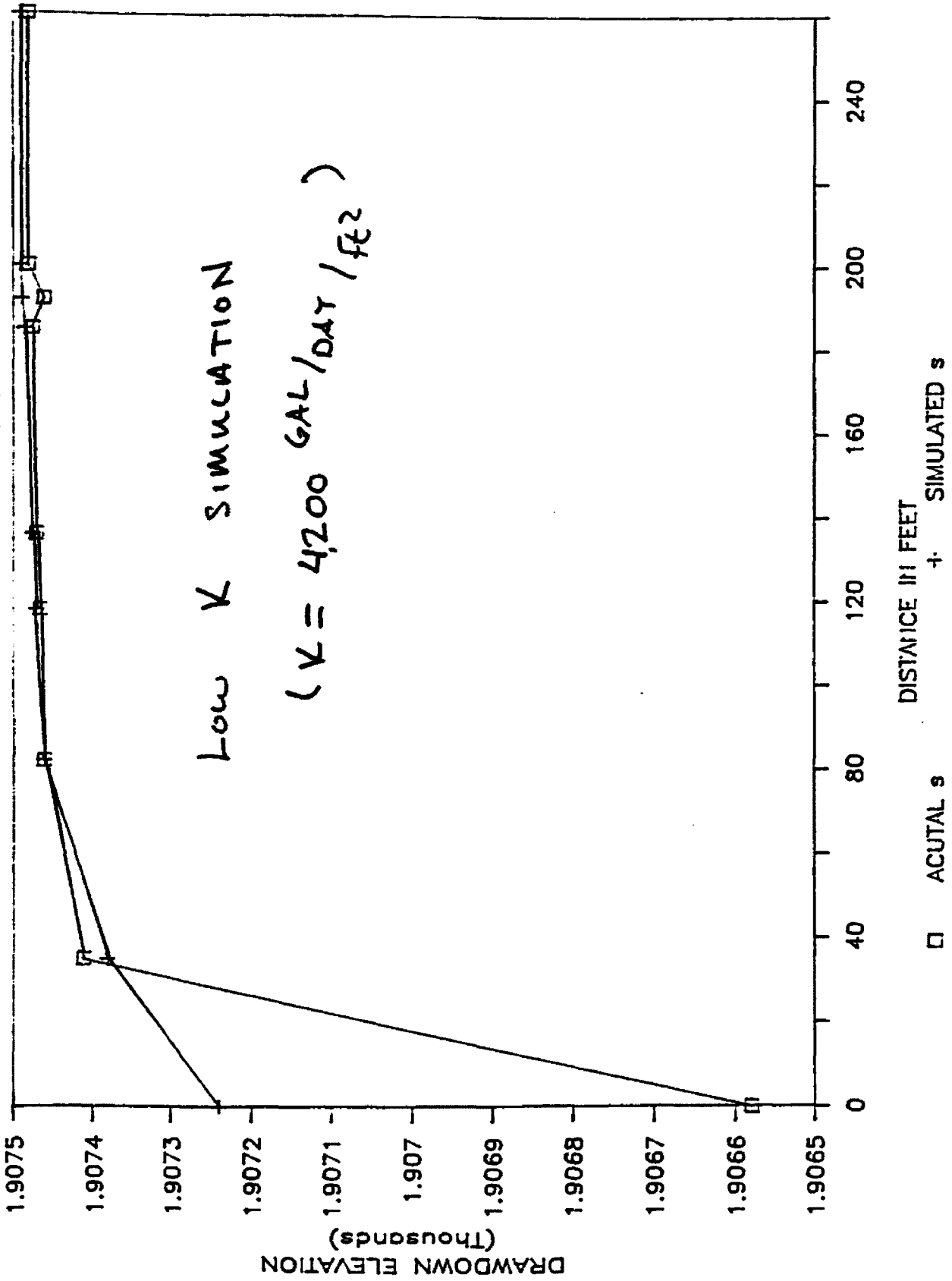
ACUTAL VS. SIMULATED DRAWDOWN

PUMPING = 30 GPM, T = 192500 G/D/ft



ACUTAL VS. SIMULATED DRAWDOWN

PUMPING = 30 GPM, T = 77000 G/D/ft



COMPARISON OF DRAWDOWN OF ACT. VS. SIM. AT
TWO DIFFERENT PUMPING NODES AND TWO K VALUES

DISTANCE AWAY	ACTUAL s	PUMP=	PUMP=	PUMP=	PUMP=
		14,11 LOW K	13,10 LOW K	13,10 HI K	14,11 HI K
0	1906.58	1907.24	1907.28	1907.4	1907.41
35	1907.41	1907.42	1907.41	1907.46	1907.46
82	1907.46	1907.44	1907.44	1907.48	1907.48
118	1907.465	1907.47	1907.46	1907.49	1907.49
136	1907.47	1907.48	1907.48	1907.495	1907.49
185	1907.475	1907.48	1907.48	1907.5	1907.495
192	1907.46	1907.49	1907.48		1907.495
200	1907.48	1907.49	1907.48		1907.5
260	1907.48	1907.49	1907.47		

DATA FOR MODEL RUN WITH K = 11000 AND T = 192500.

PUMPING NODE 13,10			PUMPING NODE 14,11		
DISTANCE	DRAWDOWN	WL ELV.	DISTANCE	DRAWDOWN	WL ELV.
0	0.1	1907.4	0	0.09	1907.41
35	0.04	1907.46	35	0.04	1907.46
82	0.02	1907.48	82	0.02	1907.48
118	0.01	1907.49	118	0.01	1907.49
136	0.005	1907.495	136	0.01	1907.49
185	0	1907.5	185	0.005	1907.495
			192	0.005	1907.495
			200	0	1907.5

DATA FOR MODEL RUN WITH K = 4400 AND T = 77000.

PUMPING NODE 13,10			PUMPING NODE 14,11		
DISTANCE	DRAWDOWN	WL ELV.	DISTANCE	DRAWDOWN	WL ELV.
0	0.22	1907.28	0	0.26	1907.24
35	0.09	1907.41	40	0.08	1907.42
82	0.06	1907.44	80	0.06	1907.44
118	0.04	1907.46	120	0.03	1907.47
136	0.02	1907.48	160	0.02	1907.48
185	0.02	1907.48	200	0.02	1907.48
192	0.02	1907.48	240	0.01	1907.49
200	0.02	1907.48	280	0.01	1907.49
260	0.03	1907.47	320	0.01	1907.49

ACTUAL AQ. TEST DATA		
DISTANCE	DRAWDOWN	WL ELV.
0	0.92	1906.58
35	0.09	1907.41
82	0.04	1907.46
118	0.035	1907.465
136	0.03	1907.47
185	0.025	1907.475
192	0.04	1907.46
200	0.02	1907.48
260	0.02	1907.48

STEADY-STATE STATISTICS (OCT.)

HEAD			
WELL	ACTUAL	SIMULATED	DIFF.
2	1908.2	1908.5	-0.3
3	1908.1	1908.26	-0.16
4	1907.93	1907.91	0.02
5	1907.85	1907.85	0
6	1907.76	1907.7	0.06
7	1907.6	1907.5	0.1
8	1906.93	1906.81	0.12
11	1907.68	1907.75	-0.07
12	1907.92	1907.97	-0.05
13	1907.08	1907.05	0.03
14	1908.12	1908.05	0.07
15	1907.76	1907.74	0.02
16	1907.7	1907.7	0
17	1907.32	1907.24	0.08
18	1907.2	1907.13	0.07
19	1907.68	1907.64	0.04
20	1907.72	1907.7	0.02
21	1908.18	1908.5	-0.32

HEAD			
WELL	ACTUAL	SIMULATED	DIFF.
3	1908.99	1908.93	0.06
4	1908.78	1908.56	0.22
5	1908.69	1908.48	0.21
7	1908.38	1908.27	0.11
6	1908.56	1908.47	0.09
14	1908.57	1908.8	-0.23
OCT 2	1908.2	1908.5	-0.3
3	1908.1	1908.26	-0.16
4	1907.93	1907.91	0.02
5	1907.85	1907.85	0
6	1907.76	1907.7	0.06
7	1907.6	1907.5	0.1
8	1906.93	1906.81	0.12
11	1907.68	1907.75	-0.07
12	1907.92	1907.97	-0.05
13	1907.08	1907.05	0.03
14	1908.12	1908.05	0.07
15	1907.76	1907.74	0.02
16	1907.7	1907.7	0
17	1907.32	1907.24	0.08
18	1907.2	1907.13	0.07
19	1907.68	1907.64	0.04
20	1907.72	1907.7	0.02
21	1908.18	1908.5	-0.32

TRANSIENT STATISTICS

HEAD			
WELL	ACTUAL	SIMULATED	DIFF.
JUL 8	1908.56	1908.66	-0.1
13	1908.74	1908.98	-0.24
11	1909.38	1909.6	-0.22
12	1909.68	1909.74	-0.06
2	1910.11	1910.42	-0.31
3	1909.94	1910.12	-0.18
4	1909.9	1909.66	0.24
5	1909.65	1909.6	0.05
7	1909.35	1909.27	0.08
6	1909.55	1909.5	0.05
14	1909.96	1909.9	0.06
AUG 8	1908.79	1908.98	-0.19
13	1909	1909.41	-0.41
11	1909.76	1909.94	-0.18
12	1910.03	1910.04	-0.01
2	1910.38	1910.59	-0.21
3	1910.18	1910.39	-0.21
4	1910.13	1910.01	0.12
5	1910.01	1909.95	0.06
7	1910.68	1910.69	-0.01
6	1909.9	1909.87	0.03
14	1910.34	1910.23	0.11
SEP 13	1907.81	1907.99	-0.18
11	1908.45	1908.45	0
12	1908.23	1908.46	-0.23
2	1909.04	1909.07	-0.03

HEAD			
WELL	ACTUAL	SIMULATED	DIFF.
NOV 8	1906.31	1906.37	-0.06
13	1906.72	1906.56	0.16
18	1906.91	1906.65	0.26
11	1907.25	1907.23	0.02
12	1907.49	1907.47	0.02
2	1907.73	1908	-0.27
21	1907.72	1908.02	-0.3
3	1907.64	1907.77	-0.13
4	1907.54	1907.41	0.13
5	1907.42	1907.37	0.05
16	1907.3	1907.21	0.09
20	1907.25	1907.23	0.02
7	1907.19	1907.04	0.15
19	1907.27	1907.18	0.09
6	1907.34	1907.24	0.1
15	1907.33	1907.28	0.05
14	1907.66	1907.58	0.08
17	1906.96	1906.79	0.17
DEC 8	1906.45	1906.3	0.15
13	1906.53	1906.51	0.02
18	1906.66	1906.59	0.07
11	1907.04	1907	0.04
12	1907.26	1907.3	-0.04
2	1907.48	1907.71	-0.23

	WELL	ACUTAL	SIMULATED	DIFF.		WELL	ACUTAL	SIMULATED	DIFF.
	21	1907.46	1907.71	-0.25		12	1906.88	1906.88	0
	3	1907.39	1907.51	-0.12		2	1906.99	1907.11	-0.12
	4	1907.3	1907.21	0.09		21	1907.04	1907.09	-0.05
	5	1907.2	1907.19	0.01		4	1906.93	1906.76	0.17
	16	1907.08	1907.04	0.04		5	1906.84	1906.78	0.06
	20	1907.09	1907.06	0.03		16	1906.73	1906.68	0.05
	7	1907	1906.9	0.1		20	1906.74	1906.67	0.07
	19	1907.04	1907.01	0.03		7	1906.65	1906.66	-0.01
	6	1907.11	1907.06	0.05		19	1906.7	1906.62	0.08
	15	1907.12	1907.09	0.03		6	1906.75	1906.64	0.11
	14	1907.4	1907.33	0.07		15	1906.75	1906.66	0.09
	17	1906.64	1906.68	-0.04		14	1907.02	1906.79	0.23
JAN	8	1906.34	1906.27	0.07		17	1906.43	1906.43	0
	13	1906.44	1906.51	-0.07	APR	8	1906.11	1905.7	0.41
	18	1906.57	1906.58	-0.01		13	1906.14	1906.17	-0.03
	11	1906.92	1906.98	-0.06		18	1906.29	1906.25	0.04
	12	1907.09	1907.17	-0.08		11	1906.58	1906.54	0.04
	2	1907.29	1907.59	-0.3		12	1906.77	1906.76	0.01
	21	1907.27	1907.58	-0.31		2	1906.95	1907.07	-0.12
	3	1907.2	1907.39	-0.19		21	1906.94	1907.05	-0.11
	4	1907.15	1907.1	0.05		3	1906.92	1906.88	0.04
	5	1907.05	1907.08	-0.03		14	1906.89	1906.68	0.21
	16	1907.83	1907.94	-0.11		5	1906.73	1906.64	0.09
	20	1906.86	1906.97	-0.11		4	1906.81	1906.63	0.18
	7	1906.86	1906.83	0.03		16	1906.62	1906.52	0.1
	19	1906.89	1906.92	-0.03		19	1906.58	1906.46	0.12
	6	1906.97	1906.97	0		6	1906.64	1906.49	0.15
	15	1906.96	1907	-0.04		15	1906.63	1906.51	0.12
	14	1907.22	1907.22	0		7	1906.55	1906.4	0.15
	17	1906.62	1906.65	-0.03		17	1906.33	1906.23	0.1
FEB	8	1906.23	1906.08	0.15	MAY	8	1908.25	1908.7	-0.45
	13	1906.3	1906.35	-0.05		13	1908.52	1909.08	-0.56
	18	1906.46	1906.43	0.03		18	1908.77	1909.18	-0.41
	11	1906.79	1906.78	0.01		11	1909.4	1909.68	-0.28
	12	1906.97	1906.96	0.01		12	1909.96	1909.76	0.2
	2	1907.17	1907.34	-0.17		2	1910.29	1910.45	-0.16
	3	1907.08	1907.17	-0.09		21	1910.27	1910.53	-0.26
	4	1907.02	1906.9	0.12		3	1910.16	1910.23	-0.07
	5	1906.9	1906.88	0.02		14	1910.2	1910.05	0.15
	16	1906.83	1906.76	0.07		5	1909.86	1909.74	0.12
	20	1906.83	1906.77	0.06		4	1910	1909.8	0.2
	7	1906.75	1906.65	0.1		16	1909.65	1909.59	0.06
	19	1906.78	1906.73	0.05		19	1909.6	1909.58	0.02
	6	1906.88	1906.77	0.11		6	1909.7	1909.65	0.05
	15	1906.81	1906.8	0.01		15	1909.75	1909.69	0.06
	17	1906.5	1906.48	0.02		7	1909.46	1909.46	0
MAR	8	1906.17	1905.99	0.18		17	1909.28	1909.02	0.26
	13	1906.23	1906.38	-0.15	JUN	13	1908.8	1908.92	-0.12
	18	1906.39	1906.44	-0.05		18	1908.79	1909.01	-0.22
	11	1906.7	1906.69	0.01		11	1909.58	1909.55	0.03

WELL	ACTUAL	SIMULATED	DIFF.
12	1909.86	1909.57	0.29
2	1910.41	1910.34	0.07
21	1910.44	1910.49	-0.05
3	1910.25	1910.15	0.1
14	1910.34	1910.01	0.33
5	1909.74	1909.57	0.17
4	1909.96	1909.69	0.27
16	1909.51	1909.44	0.07
19	1909.55	1909.47	0.08
6	1909.67	1909.55	0.12
15	1909.73	1909.58	0.15
7	1909.37	1909.31	0.06
17	1908.92	1909.12	-0.2

STEADY-STATE RESULTS

Regression Output:

Constant	0
Std Err of Y Est	0.126713
R Squared	0.926028
No. of Observations	18
Degrees of Freedom	17

X Coefficient(s) 1.000020

Std Err of Coef. 0.000015

TRANSIENT RESULTS

Regression Output:

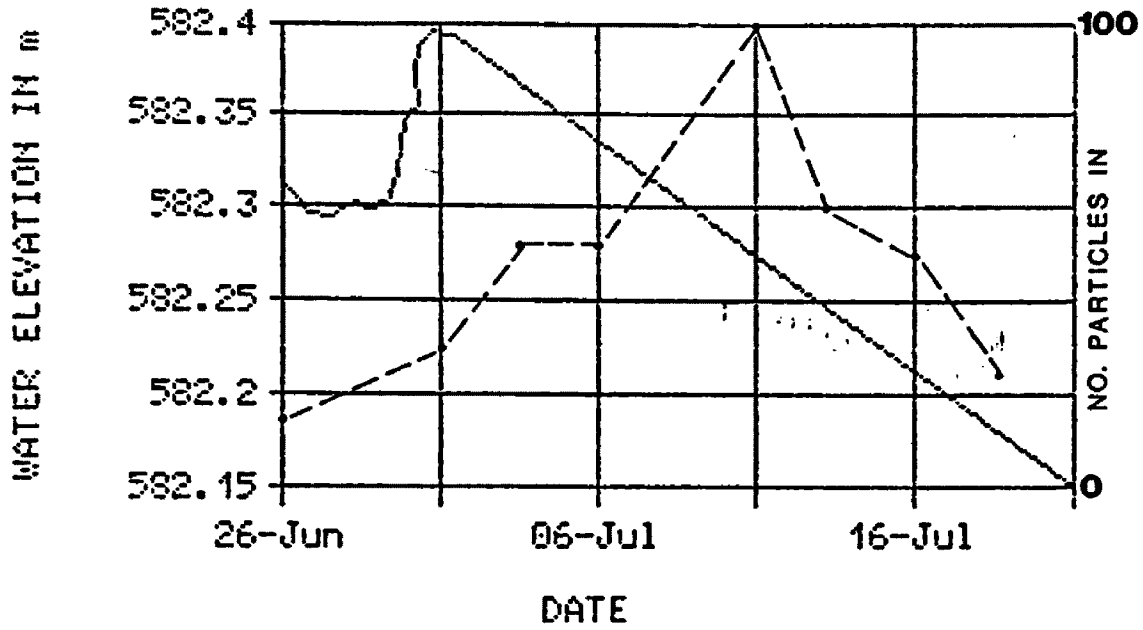
Constant	0
Std Err of Y Est	0.171654
R Squared	0.993228
No. of Observations	187
Degrees of Freedom	186

X Coefficient(s) 1.000001

Std Err of Coef. 0.000006

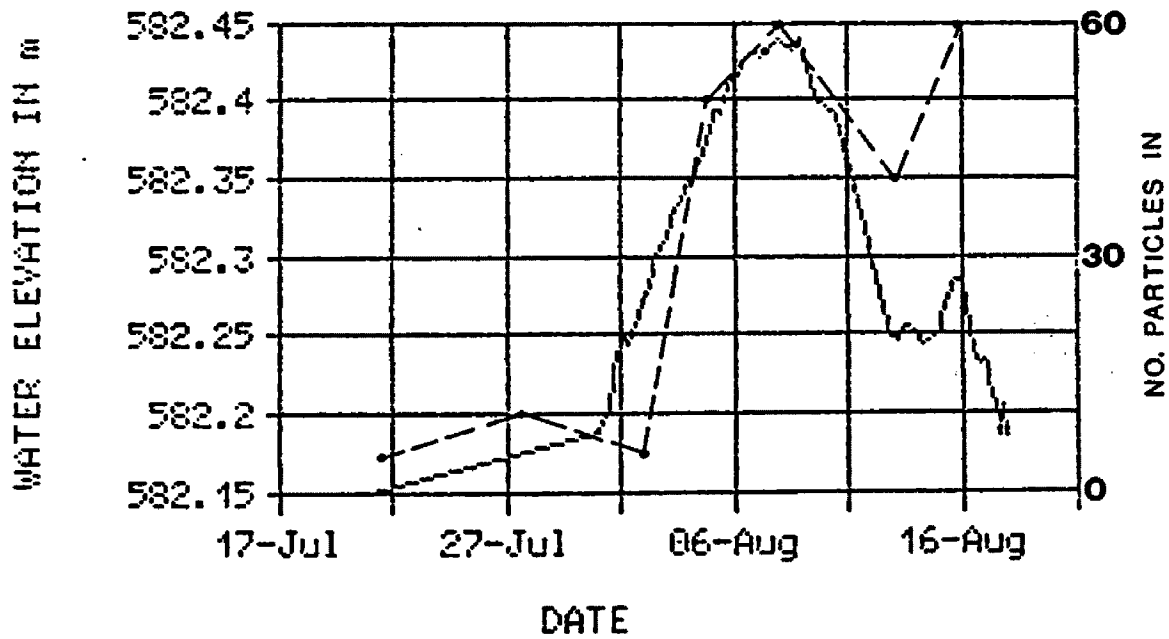
APPENDIX U
RANDOM-WALK LOADING

JULY WATER LEVEL (6/26 TO 7/21)
MW 3

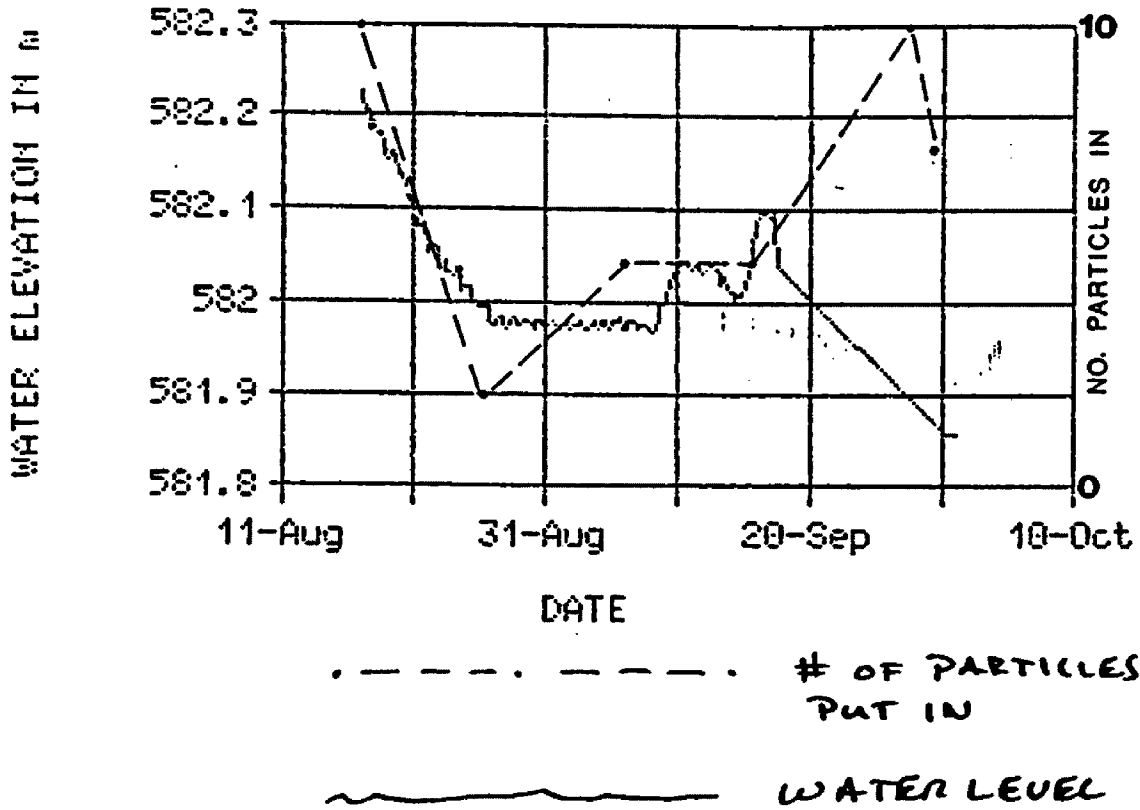


- - - - - # OF PARTICLES PUT IN.
 _____ WATER LEVEL.

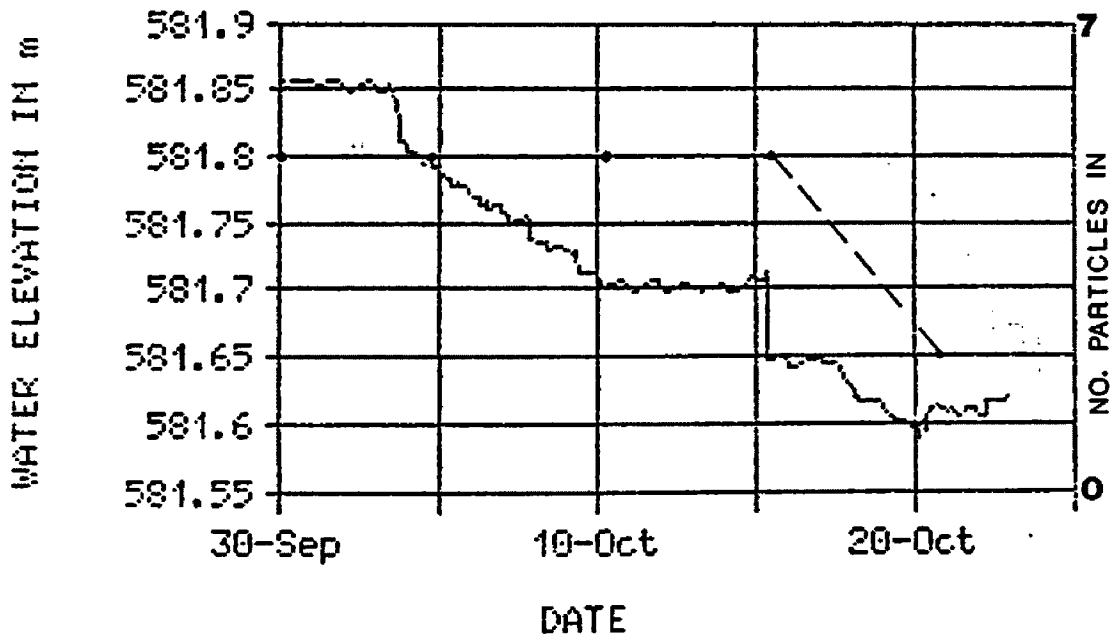
AUG. WATER LEVEL (7/21 TO 8/17)
MW 3



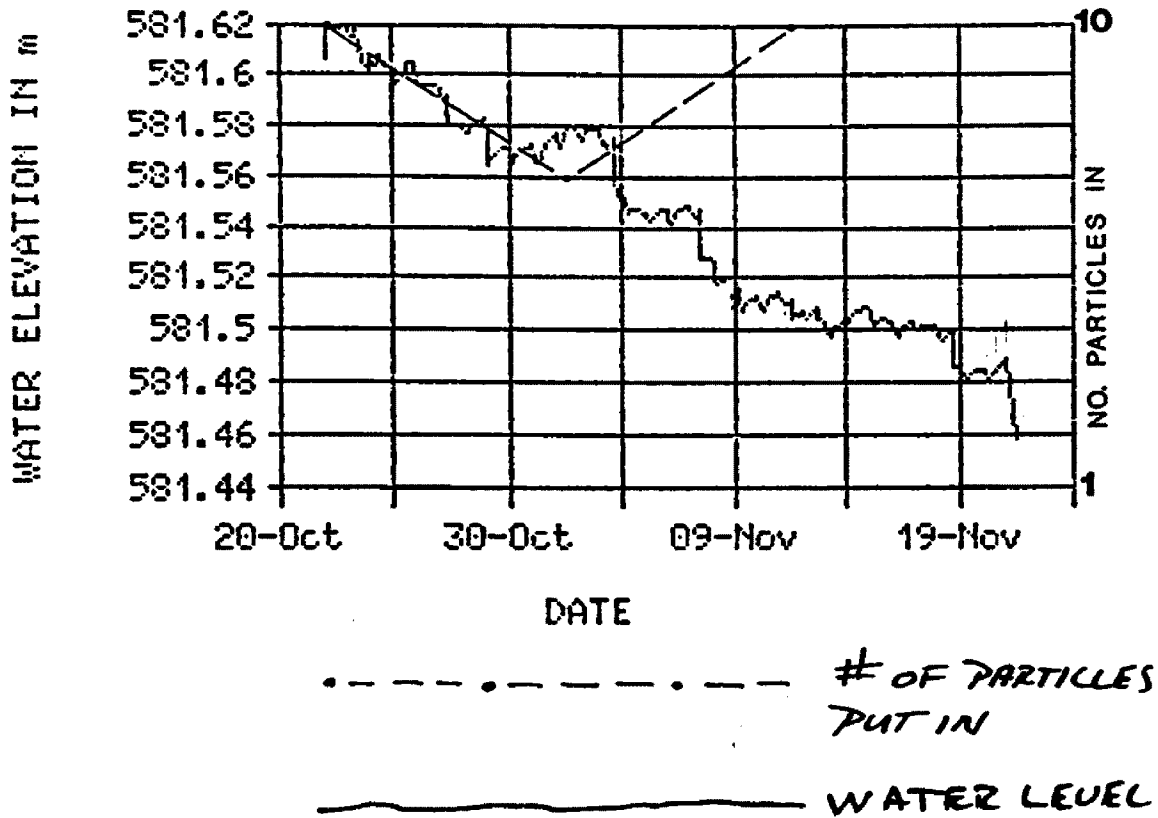
SEP. WATER LEVEL (8/17 TO 9/30)
MW 3



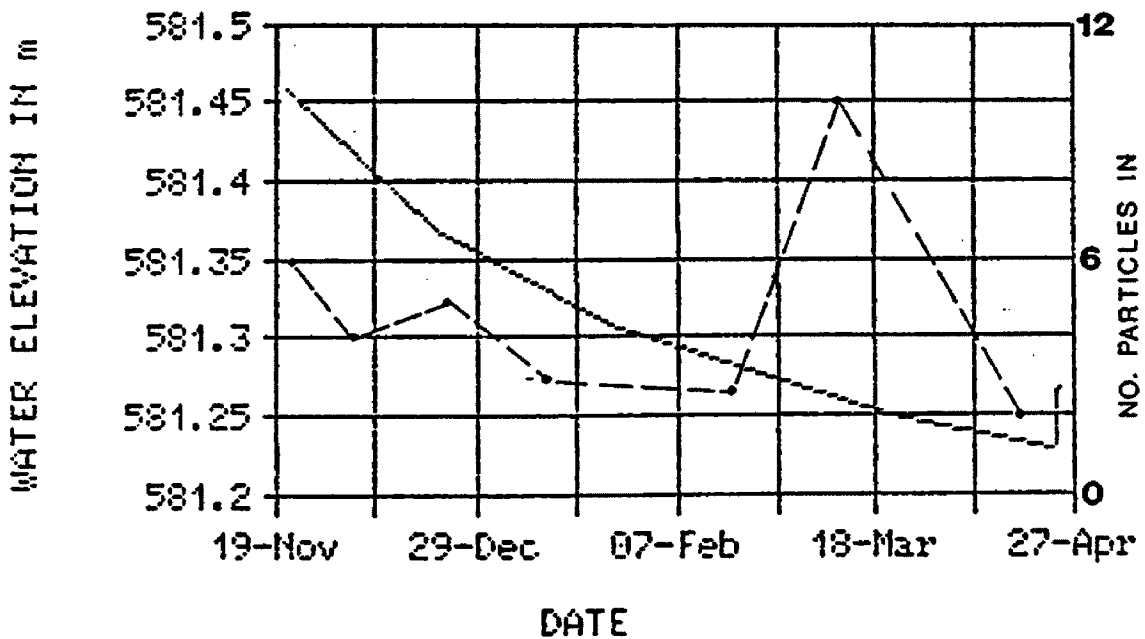
OCT. WATER LEVEL (9/30 TO 10/22)
MW 3



NOV. WATER LEVEL (10/22 TO 11/22)
MW 3

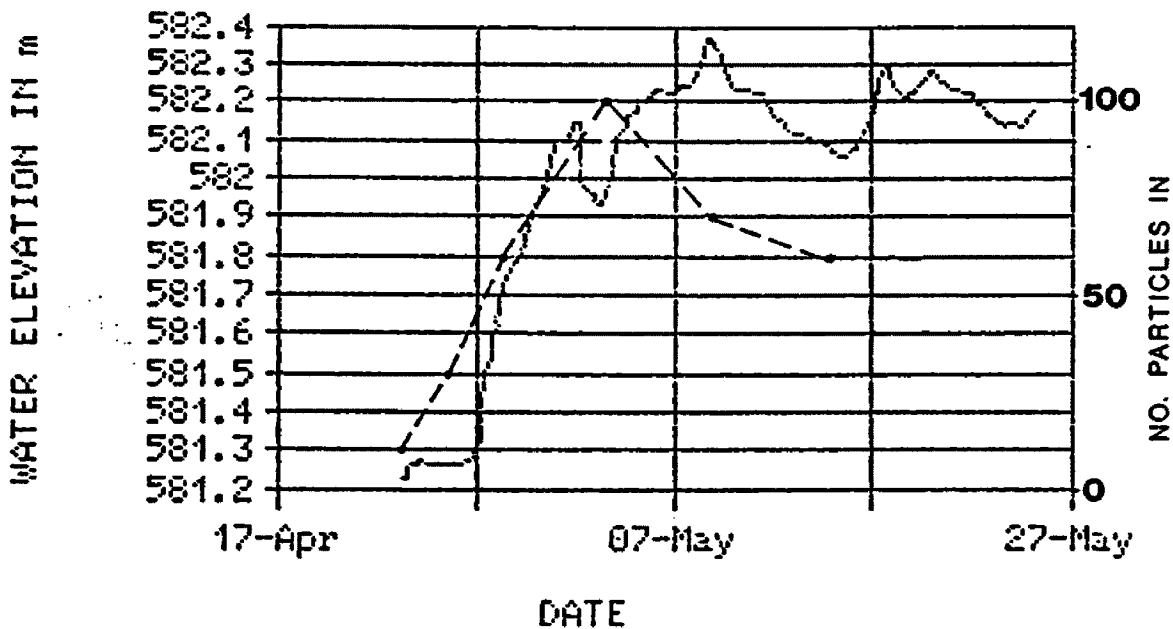


DEC.- APR. WATER LEVEL (11/22 TO 4/23)
MW 3



MAY WATER LEVEL (4/23 TO 5/24)

MW 3

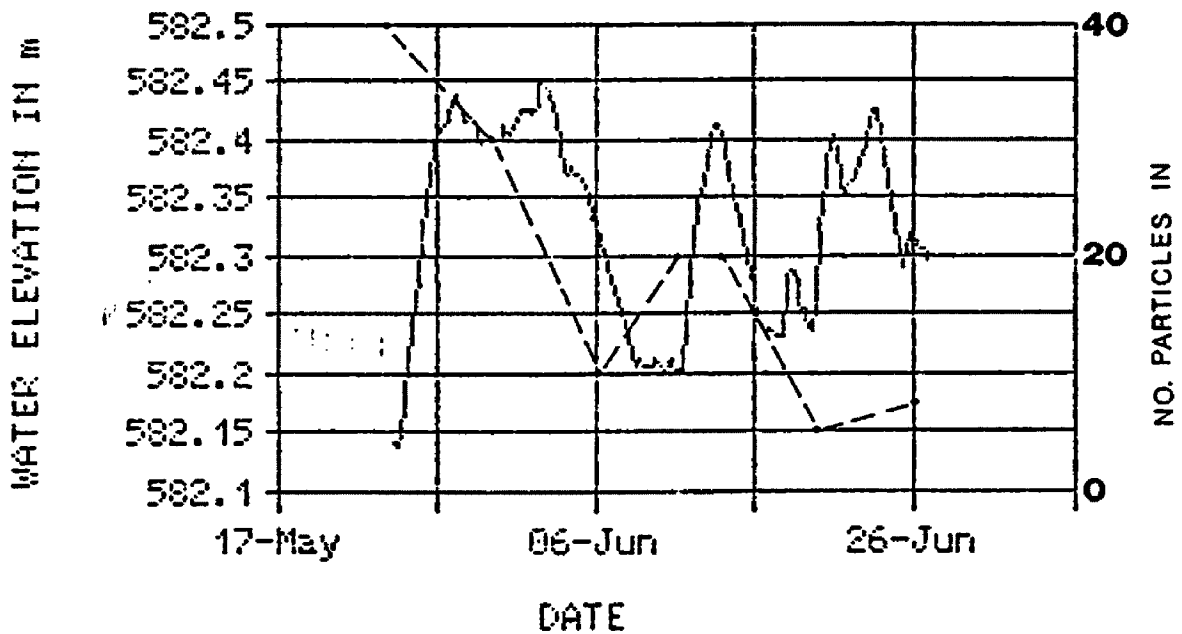


--- # OF PARTICLES PUT IN

_____ WATER LEVEL

JUNE WATER LEVEL (5/24 TO 6/26)

MW 3



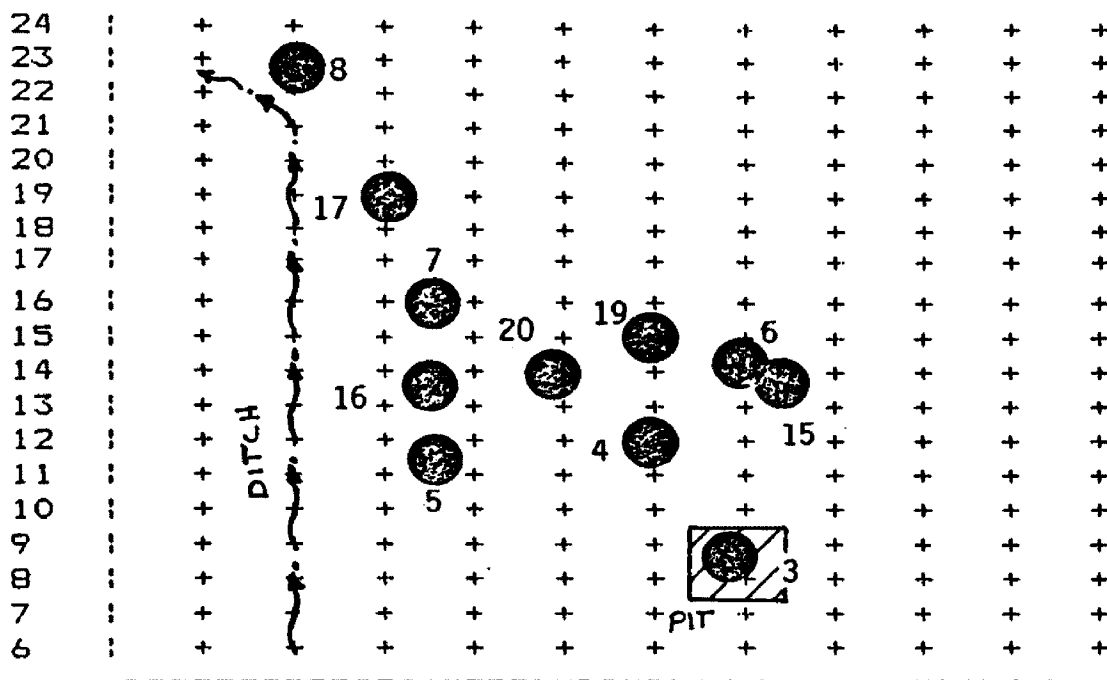
APPENDIX V

SIMULATED Cl PLOTS AND ACTUAL VERSES SIMULATED CHLORIDE DATA

ACCUMULATED TIME = 0 DAYS
 PARTICLE MAP

PARTICLES = 0

----- WELL SITE LOCATION
 COORDINATES ARE IN FEET



8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

ACCUMULATED TIME = 30 DAYS PARTICLES = 328
 CONCENTRATION MAP IN PPM

----- COORDINATES ARE IN FEET

16	:	+	+	+	+	+	+	+	+	+
15	:	+	+	+	+	+	+	+	+	+
14	:	+	+	+	127	95	+	+	+	+
13	:	+	+	+	295	473	126	42	+	+
12	:	+	+	+	+	785	563	+	+	+
11	:	+	+	+	+	625	1029	247	+	+
10	:	+	+	+	+	+	1335	743	+	+
9	:	+	+	+	+	+	463	1127	15	+
8	:	+	+	+	+	+	31	654	+	+
7	:	+	+	+	+	+	+	41	15	+
6	:	+	+	+	+	+	+	+	+	+



8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. NAY

ACCUMULATED TIME = 90 DAYS PARTICLES = 552
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	+	+	+	+	+	+	+	+	+	+
21	:	+	+	+	+	+	+	+	+	+	+	+	+	+
20	:	+	+	+	+	+	+	+	+	+	+	+	+	+
19	:	+	+	63	+	32	+	+	+	+	+	+	+	+
18	:	+	+	63	42	+	+	+	+	+	+	+	+	+
17	:	+	+	+	42	157	+	+	+	+	+	+	+	+
16	:	+	+	108	209	94	31	+	+	+	+	+	+	+
15	:	+	+	187	749	280	31	21	+	+	+	+	+	+
14	:	+	+	62	705	590	155	83	+	+	+	+	+	+
13	:	+	+	+	331	836	278	82	+	+	+	+	+	+
12	:	+	+	+	+	278	1139	184	+	+	+	+	+	+
11	:	+	+	+	41	154	1779	388	+	+	+	+	+	+
10	:	+	+	+	+	+	311	2221	220	+	+	+	+	+
9	:	+	+	+	+	+	304	2046	+	+	+	+	+	+
8	:	+	+	+	+	+	30	948	15	+	+	+	+	+
7	:	+	+	+	+	+	+	69	15	+	+	+	+	+
6	:	+	+	+	+	+	+	15	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. JUL

ACCUMULATED TIME = 120 DAYS PARTICLES = 702
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	31	+	+	+	+	+	+	+	+	+
21	:	+	+	+	+	+	+	+	+	+	+	+	+	+
20	:	+	+	+	+	+	+	+	+	+	+	+	+	+
19	:	+	+	68	45	34	+	+	+	+	+	+	+	+
18	:	+	+	+	91	102	34	+	+	+	+	+	+	+
17	:	+	+	136	588	169	67	+	+	+	+	+	+	+
16	:	+	+	203	811	405	34	+	+	+	+	+	+	+
15	:	+	+	68	1034	673	134	+	+	+	+	+	+	+
14	:	+	+	+	121	1114	368	89	+	+	+	+	+	+
13	:	+	+	+	402	1605	701	22	+	+	+	+	+	+
12	:	+	+	+	45	900	1297	285	+	+	+	+	+	+
11	:	+	+	+	44	532	1459	375	+	+	+	+	+	+
10	:	+	+	+	+	+	1322	967	+	+	+	+	+	+
9	:	+	+	+	+	+	560	2561	16	+	+	+	+	+
8	:	+	+	+	+	+	+	1134	+	+	+	+	+	+
7	:	+	+	+	+	+	+	65	32	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. AUG



ACCUMULATED TIME = 60 DAYS PARTICLES = 320
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	+	+	+	+	+	+	+	+	+	+
21	:	+	+	+	+	+	+	+	+	+	+	+	+	+
20	:	+	+	+	+	+	+	+	+	+	+	+	+	+
19	:	+	+	+	+	+	+	+	+	+	+	+	+	+
18	:	+	+	+	+	+	+	+	+	+	+	+	+	+
17	:	+	+	+	43	32	+	+	+	+	+	+	+	+
16	:	+	+	129	128	+	+	+	+	+	+	+	+	+
15	:	+	+	255	298	127	+	+	+	+	+	+	+	+
14	:	+	+	847	412	32	21	+	+	+	+	+	+	+
13	:	+	+	127	1434	726	94	21	+	+	+	+	+	+
12	:	+	+	63	882	1382	345	164	+	+	+	+	+	+
11	:	+	+	167	532	718	187	+	+	+	+	+	+	+
10	:	+	+	+	93	559	251	+	+	+	+	+	+	+
9	:	+	+	+	+	31	411	31	+	+	+	+	+	+
8	:	+	+	+	+	62	161	15	+	+	+	+	+	+
7	:	+	+	+	+	41	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	15	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. JUN

ACCUMULATED TIME = 180 DAYS PARTICLES = 300
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	+	+	+	+	+	+	+	+	+	+
21	:	+	+	+	+	+	34	+	+	+	+	+	+	+
20	:	+	+	+	91	34	+	+	+	+	+	+	+	+
19	:	+	+	68	91	+	+	+	+	+	+	+	+	+
18	:	+	+	204	91	34	34	+	+	+	+	+	+	+
17	:	+	+	68	136	169	+	+	+	+	+	+	+	+
16	:	+	+	68	541	236	34	+	+	+	+	+	+	+
15	:	+	+	135	1034	303	34	+	+	+	+	+	+	+
14	:	+	+	404	942	738	201	+	+	+	+	+	+	+
13	:	+	+	336	1341	970	167	22	+	+	+	+	+	+
12	:	+	+	134	624	1067	965	88	+	+	+	+	+	+
11	:	+	+	133	532	497	132	33	+	+	+	+	+	+
10	:	+	+	+	33	363	132	+	+	+	+	+	+	+
9	:	+	+	+	33	99	306	16	+	+	+	+	+	+
8	:	+	+	+	+	+	153	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	16	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. SEP

ACCUMULATED TIME = 180 DAYS PARTICLES = 290
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	50	+	+	+	+	+	+	+	+	+	+
21	:	+	+	74	+	+	+	+	+	+	+	+	+	+
20	:	+	+	148	37	+	+	+	+	+	+	+	+	+
19	:	+	+	148	147	+	+	+	+	+	+	+	+	+
18	:	+	+	73	391	146	37	+	+	+	+	+	+	+
17	:	+	+	146	925	146	+	24	+	+	+	+	+	+
16	:	+	+	2911	965	217	109	+	+	+	+	+	+	+
15	:	+	+	1451	301	541	144	24	+	+	+	+	+	+
14	:	+	+	216	767	574	323	72	18	+	+	+	+	+
13	:	+	+	382	500	428	71	+	+	+	+	+	+	+
12	:	+	+	190	356	462	47	18	+	+	+	+	+	+
11	:	+	+	95	71	105	188	+	+	+	+	+	+	+
10	:	+	+	+	+	+	141	211	+	+	+	+	+	+
9	:	+	+	+	+	+	+	233	+	+	+	+	+	+
8	:	+	+	+	+	+	+	66	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. OCT

ACCUMULATED TIME = 240 DAYS PARTICLES = 146
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	+	+	+	+	+	+	+	+	+	+
21	:	+	+	75	100	+	+	+	+	+	+	+	+	+
20	:	+	+	149	37	+	25	+	+	+	+	+	+	+
19	:	+	+	149	595	257	37	25	+	+	+	+	+	+
18	:	+	+	148	543	185	37	+	+	+	+	+	+	+
17	:	+	+	148	221	74	45	+	+	+	+	+	+	+
16	:	+	+	176	220	110	24	+	+	+	+	+	+	+
15	:	+	+	90	256	293	24	+	+	+	+	+	+	+
14	:	+	+	146	73	75	+	+	+	+	+	+	+	+
13	:	+	+	109	291	48	+	+	+	+	+	+	+	+
12	:	+	+	72	109	24	+	+	+	+	+	+	+	+
11	:	+	+	108	144	+	+	+	+	+	+	+	+	+
10	:	+	+	+	+	+	36	167	+	+	+	+	+	+
9	:	+	+	+	+	+	+	36	72	+	+	+	+	+
8	:	+	+	+	+	+	+	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. DEC



ACCUMULATED TIME = 210 DAYS PARTICLES = 167
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	+	+	+	+	+	+	+	+	+	+	+
21	:	+	+	50	+	+	+	+	+	+	+	+	+	+
20	:	+	+	258	+	+	+	+	+	+	+	+	+	+
19	:	+	+	149	247	111	+	25	+	+	+	+	+	+
18	:	+	+	345	184	74	+	+	+	+	+	+	+	+
17	:	+	+	74	343	331	+	24	+	+	+	+	+	+
16	:	+	+	527	549	183	97	+	+	+	+	+	+	+
15	:	+	+	389	291	169	24	+	+	+	+	+	+	+
14	:	+	+	97	508	181	40	+	+	+	+	+	+	+
13	:	+	+	96	108	253	48	+	+	+	+	+	+	+
12	:	+	+	72	288	72	+	+	+	+	+	+	+	+
11	:	+	+	36	251	72	+	+	+	+	+	+	+	+
10	:	+	+	+	+	+	143	190	+	+	+	+	+	+
9	:	+	+	+	+	+	+	142	+	+	+	+	+	+
8	:	+	+	+	+	+	+	71	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. NOV

ACCUMULATED TIME = 270 DAYS PARTICLES = 111
 CONCENTRATION MAP IN PPM

COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	+	+	+	+	+	+	+	+	+	+
22	:	+	+	68	+	25	+	+	+	+	+	+	+	+
21	:	+	+	76	252	74	+	25	+	+	+	+	+	+
20	:	+	+	151	10	38	+	+	+	+	+	+	+	+
19	:	+	+	150	50	75	75	25	+	+	+	+	+	+
18	:	+	+	75	200	299	75	25	+	+	+	+	+	+
17	:	+	+	75	147	298	112	+	+	+	+	+	+	+
16	:	+	+	99	112	74	+	+	+	+	+	+	+	+
15	:	+	+	49	111	185	74	+	+	+	+	+	+	+
14	:	+	+	74	111	74	+	+	+	+	+	+	+	+
13	:	+	+	110	37	147	+	+	+	+	+	+	+	+
12	:	+	+	70	70	+	+	+	+	+	+	+	+	+
11	:	+	+	37	37	73	+	+	+	+	+	+	+	+
10	:	+	+	+	+	+	121	+	+	+	+	+	+	+
9	:	+	+	+	+	+	72	18	+	+	+	+	+	+
8	:	+	+	+	+	+	+	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8 9 1 1 1 1 1 1 1 1 1
 0 1 2 3 4 5 6 7 8

DO A SCREEN PRINT NOW OR PRESS <RETURN> TO GO BACK TO THE MENU. JAN

ACCUMULATED TIME = 300 DAYS PARTICLES = 97
 CONCENTRATION MAP IN PPM

 COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	13	+	+	+	+	+	+	+	+	+
22	:	+	+	51	136	51	25	17	+	+	+	+	+	+
21	:	+	+	228	202	76	38	+	+	+	+	+	+	+
20	:	+	+	151	303	76	114	+	+	+	+	+	+	+
19	:	+	+	151	75	38	+	+	+	+	+	+	+	+
18	:	+	+	150	188	75	25	+	+	+	+	+	+	+
17	:	+	+	50	187	112	50	+	+	+	+	+	+	+
16	:	+	+	50	187	150	25	+	+	+	+	+	+	+
15	:	+	+	50	37	37	50	+	+	+	+	+	+	+
14	:	+	+	50	37	111	+	+	+	+	+	+	+	+
13	:	+	+	+	74	49	+	+	+	+	+	+	+	+
12	:	+	+	+	74	49	+	+	+	+	+	+	+	+
11	:	+	+	+	+	74	+	+	+	+	+	+	+	+
10	:	+	+	+	+	24	+	+	+	+	+	+	+	+
9	:	+	+	+	+	49	+	+	+	+	+	+	+	+
8	:	+	+	+	+	24	+	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

9	9	1	1	1	1	1	1	1	1	1	1	1	1	1
		0	1	2	3	4	5	6	7	8				

DO A SCREEN PRINT NOW OR PRESS :RETURN: TO GO BACK TO THE MENU. FEB

ACCUMULATED TIME = 350 DAYS PARTICLES = 102
 CONCENTRATION MAP IN PPM

 COORDINATES ARE IN FEET

24	:	+	+	+	40	20	+	+	+	+	+	+	+	+
23	:	+	+	+	176	110	68	18	+	+	+	+	+	+
22	:	+	+	+	103	101	103	34	+	+	+	+	+	+
21	:	+	+	77	103	77	115	26	+	+	+	+	+	+
20	:	+	+	+	192	26	19	+	+	+	+	+	+	+
19	:	+	+	+	51	76	76	51	+	+	+	+	+	+
18	:	+	+	+	51	76	76	25	+	+	+	+	+	+
17	:	+	+	+	76	25	+	+	+	+	+	+	+	+
16	:	+	+	+	76	76	50	+	+	+	+	+	+	+
15	:	+	+	+	50	38	50	19	+	+	+	+	+	+
14	:	+	+	+	+	25	+	+	+	+	+	+	+	+
13	:	+	+	+	37	75	+	+	+	+	+	+	+	+
12	:	+	+	+	+	99	+	+	+	+	+	+	+	+
11	:	+	+	+	+	25	+	+	+	+	+	+	+	+
10	:	+	+	+	+	25	+	+	+	+	+	+	+	+
9	:	+	+	+	+	+	+	+	+	+	+	+	+	+
8	:	+	+	+	+	+	24	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

9	9	1	1	1	1	1	1	1	1	1	1	1	1	1
		0	1	2	3	4	5	6	7	8				

DO A SCREEN PRINT NOW OR PRESS :RETURN: TO GO BACK TO THE MENU. APR



ACCUMULATED TIME = 330 DAYS PARTICLES = 101
 CONCENTRATION MAP IN PPM

 COORDINATES ARE IN FEET

24	:	+	+	+	+	+	+	+	+	+	+	+	+	+
23	:	+	+	+	69	39	+	9	+	+	+	+	+	+
22	:	+	+	+	204	127	51	+	+	+	+	+	+	+
21	:	+	+	76	101	228	152	+	+	+	+	+	+	+
20	:	+	+	+	50	114	38	25	+	+	+	+	+	+
19	:	+	+	+	151	113	38	25	+	+	+	+	+	+
18	:	+	+	+	50	263	301	25	+	+	+	+	+	+
17	:	+	+	+	75	150	75	+	+	+	+	+	+	+
16	:	+	+	+	37	+	25	+	+	+	+	+	+	+
15	:	+	+	+	75	25	+	+	+	+	+	+	+	+
14	:	+	+	+	37	111	50	+	+	+	+	+	+	+
13	:	+	+	+	+	25	+	+	+	+	+	+	+	+
12	:	+	+	+	+	74	+	+	+	+	+	+	+	+
11	:	+	+	+	+	123	+	+	+	+	+	+	+	+
10	:	+	+	+	+	99	+	+	+	+	+	+	+	+
9	:	+	+	+	+	49	+	+	+	+	+	+	+	+
8	:	+	+	+	+	24	+	+	+	+	+	+	+	+
7	:	+	+	+	+	+	+	+	+	+	+	+	+	+
6	:	+	+	+	+	+	+	+	+	+	+	+	+	+

8	9	1	1	1	1	1	1	1	1	1	1	1	1	1
		0	1	2	3	4	5	6	7	8				

DO A SCREEN PRINT NOW OR PRESS :RETURN: TO GO BACK TO THE MENU. MAP

ACT. VS. SIM. Cl CONC. IN ppm WITH RANDOM-WALK*
 (EMPTY COLUMN SPACES REPRESENT NO DATA COLLECTION)

LOC.	MAY		JUNE		JULY	
	ACT.	SIM.	ACT.	SIM.	ACT.	SIM.
MW3	1144	1124	391	411	2640	2078
MW4	638	560	660	344	1040	1160
MW5	14	41	64	127	100**	150
MW6		0	<50	15	17	21
MW7	0	0	30	128	152	200
MW8	<34	0	<50	0	<36	0
MW16	200	126	373	255		500
MW17	<34	0	0	0		65
MW19	0	0	27	21		32
MW20		94		411		602

LOC.	AUG.		SEP.		OCT.	
	ACT.	SIM.	ACT.	SIM.	ACT.	SIM.
MW3	2764	2370		306	201	227
MW4	1254	1201		865	658	449
MW5	225**	410		275	166	185
MW6	<36	83		16		23
MW7	123	188		200	204	283
MW8	16	0		0		0
MW16		560		700	606	475
MW17		63		68	99	144
MW19		124		34	68	140
MW20		1056		738	626	559

LOC.	NOV.		DEC.		JAN.	
	ACT.	SIM.	ACT.	SIM.	ACT.	SIM.
MW3	100**	140		71		72
MW4	400**	284	300**	108	200**	72
MW5	150	95		36		36
MW6	<34	24		24		73
MW7	150	250	132	195	41	98
MW8		0		0		0
MW16	779	96		97	683	109
MW17	122	148	34	149		149
MW19	61	108	21	125		100
MW20	527	503	405	145	405	109

LOC.	FEB.		MAR.		APR.	
	ACT.	SIM.	ACT.	SIM.	ACT.	SIM.
MW3	14	48	40	49	4	24
MW4	21	73	29	70	8.2	25
MW5	21	24		0	110	0
MW6		49		25	14	50
MW7	81	50	77	37	51	50
MW8	27(?)	0	30	20(?)	11.4	30(?)
MW16	527	49	398	24	248	37
MW17	91	150	91	50	51	51
MW19	34	37	29	75	29	38
MW20	345	49	257	91	200	37

*ALL VALUES USED IN ACT. COLUMNS ARE ESTIMATED BY SUBTRACTING
20 ppm FROM CONC. AS BACKGROUND AND ANY RESIDUAL C1 CONC.
**ESTIMATES OF C1 CONC. IN SHALLOW NESTED WELLS FROM DEEP
NESTED WELL CONC.

APPENDIX W
C1 LOAD METHODOLOGY AND CONCENTRATIONS

METHODOLOGY

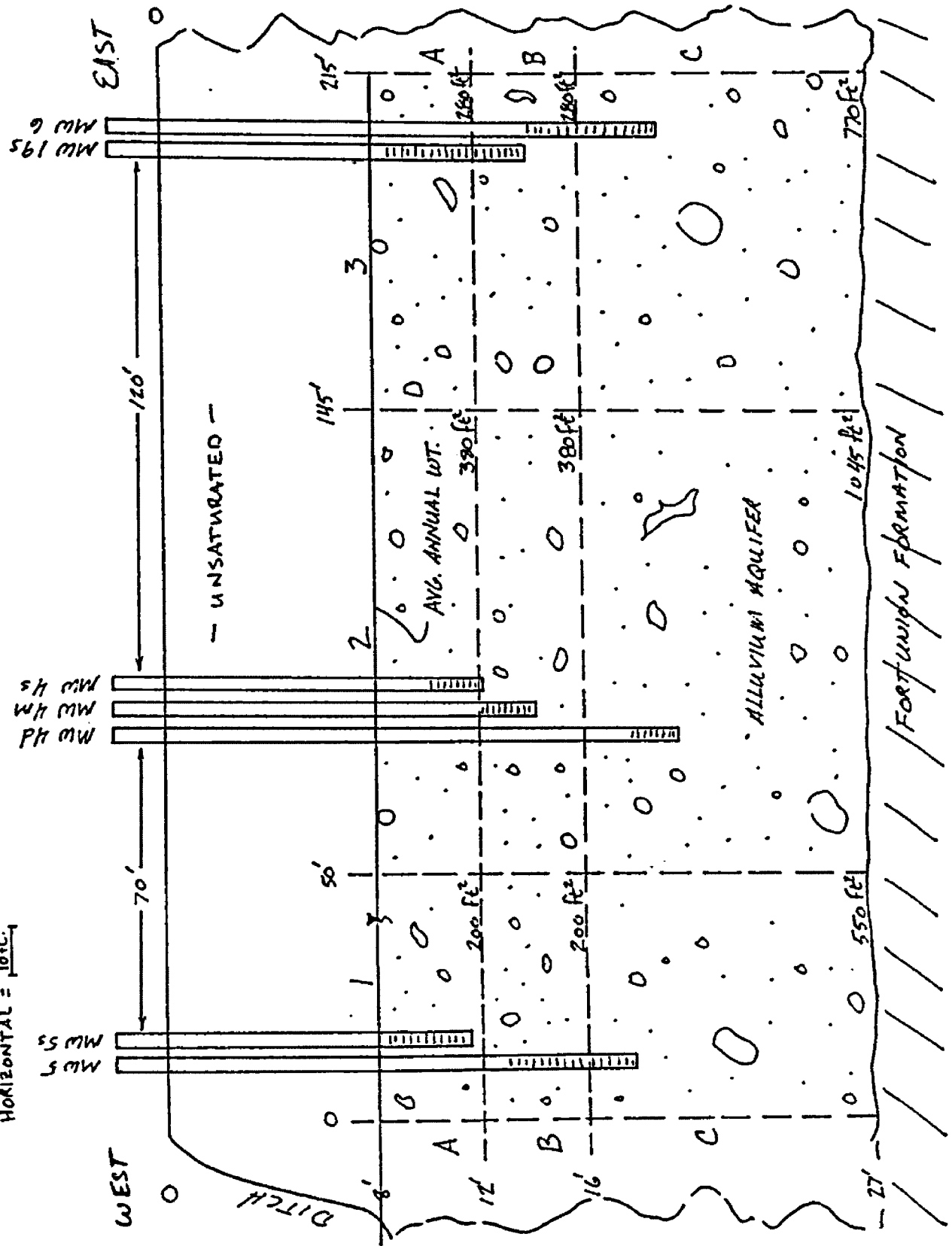
To calculate Cl loading from the reserve pit, I separated each well site with nine discrete blocks of aquifer as shown in Exhibit 12. I used data from well nests MW 4 and MW 5 to stratigraphically separate layering of Cl concentrations. Well MW 6 and MW 19s were assumed to be located at the same place for deep and shallow Cl concentration data.

I calculated loading as follows:

1. (AREA OF EACH ZONE) x (.25 POROSITY) = AREA OF WATER
(ft²)
2. (AREA OF WATER) x (MONTHLY VELOCITY) = MONTHLY WATER
VOLUME PER ZONE
(ft³)
3. CONVERT ft³ INTO LITERS
4. (MONTHLY WATER VOLUME PER ZONE) x (MONTHLY Cl mg/l) = MONTHLY
Cl LOAD (mg Cl)
5. (ADD EACH BLOCK OF Cl LOAD PER MONTH) = TOTAL MONTHLY Cl
LOAD (mg Cl)
6. (ADD EACH TOTAL MONTHLY Cl LOAD) = TOTAL YEARLY Cl
LOAD (mg CL)
7. SUBTRACT BACKGROUND Cl LEVELS OF 20 mg/l FROM TOTAL
8. CONVERT mg OF Cl LOAD PER YEAR INTO kg AND lbs OF Cl

CHLORIDE LOAD CROSS SECTION

SCALE: VERTICAL = 1 ft.
 HORIZONTAL = 10 ft.



<u>BLOCK SITE</u>	<u>CROSS SECTION DATA</u> <u>WELL DATA CHEMISTRY</u>	<u>BLOCK AREA (ft²)</u>
A1	MW 5s	200
A2	MW 4s	380
A2	MW 19s	280
B1	MW 5	200
B2	MW 4m	380
B3	MW 6	280
C1	MW 5/BACKGROUND	550
C2	MW 4d	1045
C3	MW 6/BACKGROUND	770

GROUND WATER VELOCITY DATA
MONTH VELOCITY (ft/day)

JUL	5.6
AUG	5.3
SEP	4.7
OCT	4.5
NOV	3.2
DEC	2.9
JAN	2.9
FEB	2.8
MAR	2.6
APR	2.4
MAY	5.6
JUN	5.9

C1 CONCENTRATIONS (* = ESTIMATED)

<u>BLOCK</u>	<u>JUL</u>	<u>AUG</u>	<u>SEP</u>	<u>OCT</u>
A1	225*	300*	250*	186
A2	1060	1270	1000*	678
A3	120*	100*	95*	88
B1	30	52	45*	40*
B2	650	895	400*	88
B3	37	30	28*	24*
C1	20	30	28*	25*
C2	106	36	36*	35*
C3	30	25	20*	20*

CONT. C1 CONCENTRATIONS

<u>BLOCK</u>	<u>NOV</u>	<u>DEC</u>	<u>JAN</u>	<u>FEB</u>
A1	170	150*	140*	120*
A2	400*	300*	200*	100*
A3	81	41	40*	54*
B1	34	30*	35*	41
B2	61	47	41	41
B3	22	20*	21*	21*
C1	24	22	20*	20*
C2	34	30*	30*	30*
C3	20	20*	20*	20*

<u>BLOCK</u>	<u>MAR</u>	<u>APR</u>	<u>MAY</u>	<u>JUN</u>
A1	100*	130	34	84
A2	100*	80*	658	890
A3	54	50	47	56
B1	50	60	50*	45
B2	49	28	30*	40*
B3	25*	34*	30*	40*
C1	22*	25	25*	30*
C2	30*	25	25*	35*
C3	21*	22*	25*	28*