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POTENTIAL SITE PRODUCTIVITY INFLUENCES THE RATE OF FOREST STRUCTURAL DEVELOPMENT

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Abstract. Development and maintenance of structurally complex forests in landscapes formerly managed for timber production is an increasingly common management objective. It has been postulated that the rate of forest structural development increases with site productivity. We tested this hypothesis for Douglas-fir (*Pseudotsuga menziesii* (Mirb.) Franco) forests using a network of permanent study plots established following complete timber harvest of the original old-growth forests. Forest structural development was assessed by comparing empirical measures of live tree structure to published values for Douglas-fir forests spanning a range of ages and structural conditions. The rate of forest structural development—resilience—exhibited a positive relationship with site index, a measure of potential site productivity. Density of shade-intolerant conifers declined in all study stands from an initial range of 336–4068 trees/ha to a range of 168–642 trees/ha at the most recent measurement. Angiosperm tree species declined from an initial range of 40–371 trees/ha to zero in seven of the nine plots in which they were present. Trends in shade-tolerant tree density were complex: density ranged from 0 to 575 trees/ha at the first measurement and was still highly variable (25–389 trees/ha) at the most recent measurement. Multivariate analysis identified the abundance of hardwood tree species as the strongest compositional trend apparent over the study period. However, structural variables showed a strong positive association with increasing shade-tolerant basal area and little or no association with abundance of hardwood species. Thus, while tree species succession and forest structural development occur contemporaneously, they are not equivalent processes, and their respective rates are not necessarily linearly related. The results of this study support the idea that silvicultural treatments to accelerate forest structural development should be concentrated on lower productivity sites when the management objective is reserve-wide coverage of structurally complex forests. Alternatively, high-productivity sites should be prioritized for restoration treatments when the management objective is to develop structurally complex forests on a portion of the landscape.

Key words: Douglas-fir; forest restoration; forest structural development; late-successional; productivity; *Pseudotsuga menziesii*; resilience; succession.

INTRODUCTION

Promoting the development of structurally complex forests is a primary strategy for conserving biodiversity in temperate forests (Lindenmayer and Franklin 2002, Lindenmayer et al. 2006) and an increasingly common management objective on public and some private lands (e.g., USDA and USDI 1994, City of Seattle 2000, Jenkins et al. 2004). Silvicultural treatments are one method used to promote the development of structural complexity in young, previously harvested forests (Carey 2003). Managers are faced with the challenge of prioritizing candidate forest stands for silvicultural treatment, in addition to designing site-specific treatments. Forest structure develops through a generally predictable sequence, gradually attaining complex con-

ditions many decades after high severity disturbance (Carey and Curtis 1996, Oliver and Larson 1996, Franklin and Van Pelt 2004, Spies 2004). Resilience describes the rate at which a community returns to pre-disturbance conditions (Halpern 1988); resilience is used here specifically with reference to the rate of forest structural development. Identifying factors that influence resilience of forests will help managers prioritize previously harvested forests for silvicultural treatments designed to accelerate the rate of forest structural development.

Understanding how development of community properties varies in response to changing abiotic or biotic context, such as a gradient of potential site productivity, is a promising line of investigation with the potential to increase the generality and predictive strength of community ecology (Agrawal et al. 2007). It has been postulated, but not empirically tested, that forests are more resilient—proceed through the structural development sequence faster—on high-productiv-

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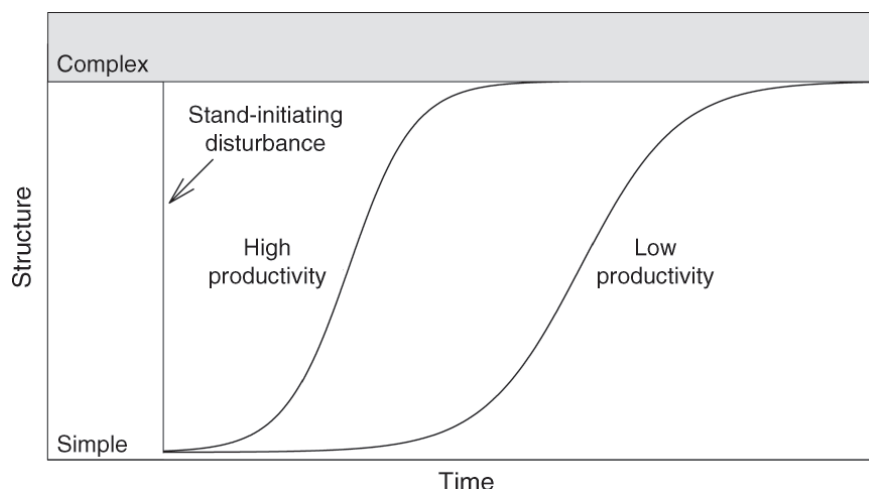


FIG. 1. Conceptual representation of the hypothesized relationship between forest structural development and site productivity. The postulated relationship is a specific case of the resilience–productivity hypothesis, which posits that resilience (defined here as the rate of forest structural development) will increase with productivity. We explicitly include anthropogenic disturbances, such as timber harvest, in this formulation of the resilience–productivity hypothesis.

ity sites (Franklin et al. 2002). This hypothesis is a specific case of the resilience–productivity hypothesis, which has seen few empirical tests (Steiner et al. 2006). Recent experimental work (Pringle et al. 2007) has caused some to suggest that low productivity ecosystems will be less able to respond to anthropogenic perturbation than their high productivity counterparts (Larson and Paine 2007). This adaptation of the resilience–productivity hypothesis could contribute to the theoretical basis for applied and restoration ecology in forest ecosystems, if upheld by empirical studies.

Succession and structural development describe different aspects of change through time in ecological communities. Succession is the process of species composition change in an ecological community following disturbance (*sensu* Connell and Slatyer 1977). Structural development describes the changes in the sizes, abundance, and spatial arrangement of the various community elements (e.g., the live trees, snags, and logs in a forest) following a disturbance (Franklin et al. 2002). Successional change of species composition and structural development are not necessarily related. No succession of tree species occurs in single-species forest systems such as edaphic climax lodgepole pine (*Pinus contorta* var. *murrayana* (Grev. and Balf.) Engelm.) forests (Franklin and Dyrness 1988). However, structural development proceeds in lodgepole pine forests as trees establish, grow, decay, and die, as mediated by interactions with and between fire, fungi, and bark beetles (Gara et al. 1985, Stuart et al. 1989). Natural single-species forests are relatively rare; in most forest systems species succession and structural development occur contemporaneously and are related.

Based on the resilience–productivity hypothesis, the development of forest structure is predicted to occur at faster rates on high-productivity sites, given similar disturbance histories (Fig. 1). We used an unplanned

experiment consisting of a set of permanent forest vegetation study plots in Douglas-fir (*Pseudotsuga menziesii* (Mirb.) Franco) forests spanning a gradient of potential site productivity to test this prediction. The permanent study plots document forest structural development following clearcut timber harvest of the original late-successional forest in the late 19th and early 20th centuries. There were three main objectives for this study: (1) conduct a test of the resilience–productivity hypothesis in the context of forest structural development; (2) characterize trends of and relationships between tree species succession and structural development; and (3) provide information that will be useful to forest managers, particularly those charged with managing for complex structure in previously harvested forests.

METHODS

Study area and field methods

The study sites are a subset of a larger group of permanent study plots (PSP), originally established as a long-term forest growth and yield study (Table 1), located within the Cedar River Municipal Watershed (CRMW) in the central western Washington Cascade Range, USA (47°24' N, 121°54' W). The original growth and yield study was designed to provide CRMW managers with an estimate of timber volume production in low elevation, second-growth Douglas-fir stands: permanent study plots were therefore located in stands spanning the productivity gradient present in the management area. The initiation of this study predates the widespread adoption of modern standards for experimental design in ecological studies: plots were subjectively located within study stands and plot sizes were determined according to the professional judgment of the responsible forester. The frequency and timing of

TABLE 1. Attributes of permanent study plots in the Cedar River Municipal Watershed, central western Washington, USA.

Plot	Size (m ²)	Site index† (m)	Elevation (m)	Year measured	Soil type‡	Soil texture‡
PSP 2	4047	27	226	1946, 1953, 1958, 1974, 1979, 1986, 2005	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 3	2023	24	486	1953, 1958, 1964, 1974, 1980, 1986, 2004	sandy-skeletal, mixed, mesic Humid Haplorthods	sandy loam
PSP 4	809	30	225	1953, 1958, 1967, 1974, 1982, 1986, 2005	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 5	809	35	270	1953, 1967, 1974, 1979, 1986, 2005	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 7	1012	38	242	1954, 1967, 1974, 1979, 1986, 2005	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 8	1012	41	244	1954, 1982, 2005	coarse-loamy, mixed, mesic, ortstein Typic Haplorthods	gravelly silt loam
PSP 9	1012	42	509	1954, 1979, 1986, 2005	coarse-loamy, mixed, mesic, ortstein Typic Haplorthods	gravelly silt loam
PSP 10	809	36	392	1954, 1967, 1974, 1986, 2003	sandy-skeletal, mixed, frigid Typic Haplorthods	gravelly sandy loam
PSP 11	1012	34	180	1954, 1974, 1979, 1986, 2005	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 12	1012	42	177	1954, 1967, 1974, 1979, 1986, 2003	sandy-skeletal, mixed, mesic Typic Haplorthods	gravelly coarse sandy loam
PSP 14	809	34	268	1954, 1974, 1981, 1986, 2006	coarse-loamy, mixed, mesic, ortstein Typic Haplorthods	sandy loam
PSP 16	809	34	294	1957, 1960, 1979, 2005	coarse-loamy, mixed, mesic, ortstein Typic Haplorthods	sandy loam

† Base age 50 years at breast height (King 1966).

‡ Goldin (1992).

remeasurements was determined by the availability of CRMW staff and funding. We analyzed the $n = 12$ plots (from a total of $n = 21$ original plots) that had not experienced silvicultural manipulation since initiation of the post-harvest cohort. All plots except one regenerated naturally; in addition to abundant natural regeneration, PSP 3 was planted with Douglas-fir seedlings.

The study plots analyzed here are distributed along a 17.4 km reach of the east-west-oriented Cedar River valley and are located within the *Tsuga heterophylla* Zone, which is typified by dominance of the shade-intolerant conifer Douglas-fir in early-successional stages (Franklin and Dyrness 1988). The shade-tolerant conifer tree species western hemlock (*Tsuga heterophylla* (Raf.) Sarg.) and western red cedar (*Thuja plicata* Donn ex D. Don) become increasingly dominant in later successional stages (Franklin et al. 2002). The climate in the Cedar River valley is characterized by rainy, cool winters and moderate, dry summers (Barg and Edmonds 1999). Average annual precipitation is 1420 mm, falling primarily as rain between October and May. The temperature regime is moderate: average annual, July and January temperatures are 9.7°, 16.9°, and 3.2°C, respectively.

Species and diameter at breast height (diameter in cm at 1.37 m above ground level; dbh) of all conifer trees present within study plots were recorded at each measurement. Diameter at breast height of all angiosperm (hardwood) trees was also measured, but species was not always recorded. The number of trees present in each PSP across all measurement periods ranged from 26 to 1935 stems per plot.

Site productivity was characterized with King's site index (King 1966). Site index was calculated using total height and age at breast height of canopy-dominant Douglas-fir; empirical age-height curves were used to determine the site index number, the height of dominant Douglas-fir at breast height age of 50 years. At least three trees were used to calculate site index for each PSP (maximum $n = 7$ trees for PSP 2).

Data analysis

Forest structure was characterized from permanent plot records for each measurement period with the old-growth index (I_{og}), a composite measure of live tree variables (Acker et al. 1998). The variable I_{og} , which is positively related to structural complexity in Douglas-fir forests (Zenner 2004), subsumes the four live-tree overstory variables identified as successfully discriminating between young, mature, and old-growth Douglas-fir stands (Spies and Franklin 1991): standard deviation of dbh for all trees >5 cm dbh; density of Douglas-fir >100 cm dbh (trees/ha); mean dbh (cm) for all trees >5 cm dbh; and density of all trees >5 cm dbh (trees/ha). Increasing values of the first three I_{og} subscores are associated with increasing old-growth structure; whereas decreasing values of the density of trees >5 cm dbh are associated with increasing old-growth structure. The I_{og} used here incorporates a slight modification of the original form and is built on more tenable assumptions about the eventual endpoint of Douglas-fir forest structural development. Specifically, the original formulation of I_{og} is based on regional mean values for old-growth structural variables. Implicit in the

original construction of I_{og} is the assumption that all forest stands converge to mean late-successional conditions. This assumption is not ideal: old-growth structure is by definition heterogeneous and many old-growth forests never achieve mean conditions. Therefore, we used the 95% confidence intervals (CI) for the regional values, rather than the mean values (Zenner 2004). I_{og} is calculated as

$$I_{og} = 25 \sum_i \left| \frac{x_i - x_{i,young}}{x_{i,old} - x_{i,young}} \right| \quad (1)$$

where x_i is the measured value for the i th structural variable; $x_{i,young}$ is the regional mean value for the i th structural variable in young Douglas-fir forests and $x_{i,old}$ is the lower bound (upper bound for density [no. trees/ha] of all trees >5 cm dbh) of the 95% CI for the i th structural variable in old-growth Douglas-fir forests. I_{og} is constrained to range from 0 to 100 (see Acker et al. 1998 for additional details); a score of 0 corresponds to typical young-forest structure while a score of 100 is interpreted as old-growth (complex) structure. In the formulation used here (Eq. 1), an I_{og} of 100 defines the minimum threshold for complex forest structure.

In order to compare resilience (i.e., the rate of forest structural development) of the different stands we fit a sigmoid curve (Spies and Franklin 1988) to the series of I_{og} measurements for each permanent plot. We fit a model of the form

$$I_{og} = \frac{100}{1 + \exp(-[t - t_{50}]/k)} \quad (2)$$

where t is stand age in years and t_{50} and k are parameters to be estimated. This particular form was desirable for several reasons. First, t_{50} is interpreted directly as an estimate of resilience, the rate of recovery toward late-successional conditions: in this case, expressed as the time in years to reach $I_{og} = 50$. Because the lower limit of I_{og} is determined by stand structure associated with forests 40–80 years old, it is expected that the I_{og} vs. age curve will be asymptotic for very young (<40 years) forests. Relative to the total duration of forest structural development (Franklin et al. 2002) rapid structural change occurs in the maturation stage (Acker et al. 1998), resulting in a steep increase in the I_{og} vs. age curve, followed by the gradual attainment of structurally complex old-growth conditions. Even seemingly catastrophic disturbances do not necessarily remove all live tree elements of forest structure (Franklin and MacMahon 2000). Thus, a flexible model that could accommodate forest ecosystem response to incomplete disturbance—a model not constrained to pass through the origin—was necessary. Curves were fit to augmented I_{og} series which included an initial values of $I_{og} = 0$ at stand age 0 years for each plot. This addition is appropriate since the study sites were harvested com-

pletely: no residual live trees from the original late-successional forest are apparent in the plot records.

Estimates of t_{50} were obtained for $n = 11$ plots ($P < 0.01$). A nonzero I_{og} score was attained at PSP 2 only in the most recent measurement period, which is insufficient to derive an estimate of t_{50} , thus PSP 2 could not be included in the productivity–resilience analysis. Visual examination of the data indicates that t_{50} at PSP 2 (site index [SI] = 27) will be approximately 100 years, consistent with observed relationship between t_{50} and SI for the remainder of the study plots (see *Results*).

Nonmetric multidimensional scaling (NMS) (Kruskal 1964) was used to characterize changes in species composition over successional time. Input variables for the NMS ordination of tree species composition were two measures of abundance (basal area [m²/ha] and stem density [no. trees/ha]) for angiosperm species, shade-tolerant conifer species, and shade-intolerant species at each plot–year combination. Stem density and basal area together provide a more complete picture of tree species composition because stem density alone does not reflect the influence of tree size on stand-scale species composition. Sorensen’s distance measure was used after data were relativized by column totals. The “slow and thorough” autopilot mode of PC-ORD version 5.0 (MjM Software Design, Gleneden Beach, Oregon, USA) in which 250 runs with the real data were analyzed for one- to six-dimensional solutions was used. A Monte Carlo test with 250 runs of randomized data indicated that the NMS ordination extracted stronger axes than expected by chance ($P < 0.01$). A final two-dimensional solution with a stress of 11.05 and instability of 1.0×10^{-7} was selected after inspecting stress values at each dimension. Pearson correlation coefficients (r) were used to characterize the relationship between the NMS axes (species-succession space) and a second matrix of structural variables comprised of the four input variables for I_{og} , as well as the calculated I_{og} scores.

RESULTS

The live tree component of forest structure developed increasing similarity to late-successional conditions on all plots over the period of observation. The density of live trees >5 cm dbh declined, while mean and standard deviation of tree diameter increased (Table 2). The coefficient of variation of tree diameter increased in all but one stand, PSP 5, over the course of the study. Two stands (PSP 8 and PSP 12) also included large diameter Douglas-fir (>100 cm dbh) by the last measurement. As of the most recent measurement I_{og} scores ranged from 20.2 to 88.0 (Fig. 2). The parameter t_{50} decreased with increasing site productivity (Fig. 3, adjusted correlation [r_{adj}^2] = 0.32, $F_{1,9} = 5.73$, $P = 0.04$), marking a relationship between resilience and productivity.

The four component variables of I_{og} did not increase uniformly during forest structural development (Fig. 4). Mean tree dbh drove initial increases in I_{og} values. Total

TABLE 2. Developmental trends of live tree structural variables in previously harvested Douglas-fir forests located in the Cedar River Municipal Watershed, western Washington Cascade Range, USA.

Site and stand age (years)	dbh of all trees >5 cm dbh (cm)		Tree density (no./ha)	
	Mean	SD	All trees >5 cm dbh	Douglas-fir >100 cm dbh
PSP 2				
24	7.3	2.5	4779	0
31	9.2	3.7	4545	0
36	10.9	4.2	3475	0
52	16.2	5.5	1934	0
57	17.9	6.2	1687	0
64	20.0	7.1	1368	0
83	20.8	12.1	595	0
PSP 3				
28	13.6	5.4	2100	0
33	15.4	6.7	1970	0
39	17.1	8.3	1775	0
49	21.9	9.0	1270	0
55	25.0	9.9	1055	0
61	27.3	11.2	950	0
79	27.7	16.3	935	0
PSP 4				
21	13.9	3.4	1247	0
26	16.3	3.9	1247	0
35	19.9	5.6	1309	0
42	23.5	6.0	1186	0
50	25.4	6.3	1074	0
54	27.2	6.9	790	0
73	29.6	11.8	692	0
PSP 5				
30	18.9	8.1	1186	0
44	27.3	8.5	865	0
51	29.2	9.5	729	0
56	31.7	8.9	716	0
63	33.5	9.7	655	0
82	41.4	10.2	457	0
PSP 7				
41	26.1	8.7	850	0
54	30.7	11.7	642	0
61	32.9	13.6	524	0
66	36.2	13.5	484	0
73	37.9	17.1	484	0
92	38.5	22.5	445	0
PSP 8				
61	40.3	13.1	514	0
89	50.1	17.8	336	0
112	31.6	29.0	524	10
PSP 9				
44	34.8	13.2	337	0
69	45.5	20.6	338	0
95	57.1	23.5	257	0
PSP 10				
27	17.1	6.6	1494	0
40	22.9	10.2	1210	0
47	28.1	12.1	976	0
59	29.2	15.2	790	0
76	33.2	17.5	630	0
PSP 11				
81	34.9	11.3	603	0
101	30.7	18.2	850	0
106	33.2	17.8	692	0
113	32.1	20.4	632	0
132	36.6	22.8	504	0

TABLE 2. Continued.

Site and stand age (years)	dbh of all trees >5 cm dbh (cm)		Tree density (no./ha)	
	Mean	SD	All trees >5 cm dbh	Douglas-fir >100 cm dbh
PSP 12				
48	33.5	9.7	672	0
61	26.9	15.3	850	0
68	33.6	16.4	583	0
73	30.8	19.3	751	0
80	28.2	20.1	662	0
97	38.4	25.2	553	10
PSP 14				
27	13.2	5.8	3038	0
47	20.8	8.4	1494	0
54	22.8	9.5	1266	0
59	24.1	10.9	976	0
79	25.5	15.5	852	0
PSP 16				
22	14.4	5.7	2174	0
25	16.6	6.0	1840	0
44	25.0	8.3	963	0
70	29.1	15.8	778	0

tree density and standard deviation of tree dbh corresponded with subsequent increases in I_{og} scores. Large diameter trees contributed to increases in I_{og} only at relatively high I_{og} scores.

A successional trend in tree species composition emerged over the duration of the study. Density of shade-intolerant conifer species (Douglas-fir, Sitka spruce [*Picea sitchensis* (Bongard) Carrière], noble fir [*Abies procera* Rehd.], and western white pine [*Pinus monticola* Douglas ex D. Don]) declined in all plots over the period of study from an initial range of 336–4068 trees/ha (mean 1382 trees/ha) to a range of 168–642 trees/ha (mean 373 trees/ha) at the most recent measurement (Fig. 5). Angiosperm trees occurred in nine of the 12 study stands. Density of angiosperm trees (bitter cherry [*Prunus emarginata* (Dougl. Ex Hook.) D. Diert.], red alder [*Alnus rubra* Bong.], willow [*Salix* sp.], and bigleaf maple [*Acer macrophyllum* Pursh]) declined from an initial range of 40–371 trees/ha (mean 123 trees/ha) to zero in seven of the nine plots in which they were present. However, PSP 10 and PSP 11 contained small amounts of bigleaf maple at the most recent measurement: 12.4 trees/ha at stand age 76 and 19.9 trees/ha at 132 years, respectively. The shade-tolerant conifer component (western hemlock, western redcedar, and Pacific silver fir [*Abies amabilis* Dougl. Ex Forbes]) exhibited complex dynamics across the study stands (Fig. 6). Shade-tolerant tree density at the initial measurement varied from 0 to 575 trees/ha (mean 151 trees/ha). By the most recent measurement shade-tolerant tree density ranged from 25 to 389 trees/ha (mean 226 trees/ha). A single stand (PSP 5) has supported very low densities of shade-tolerant trees

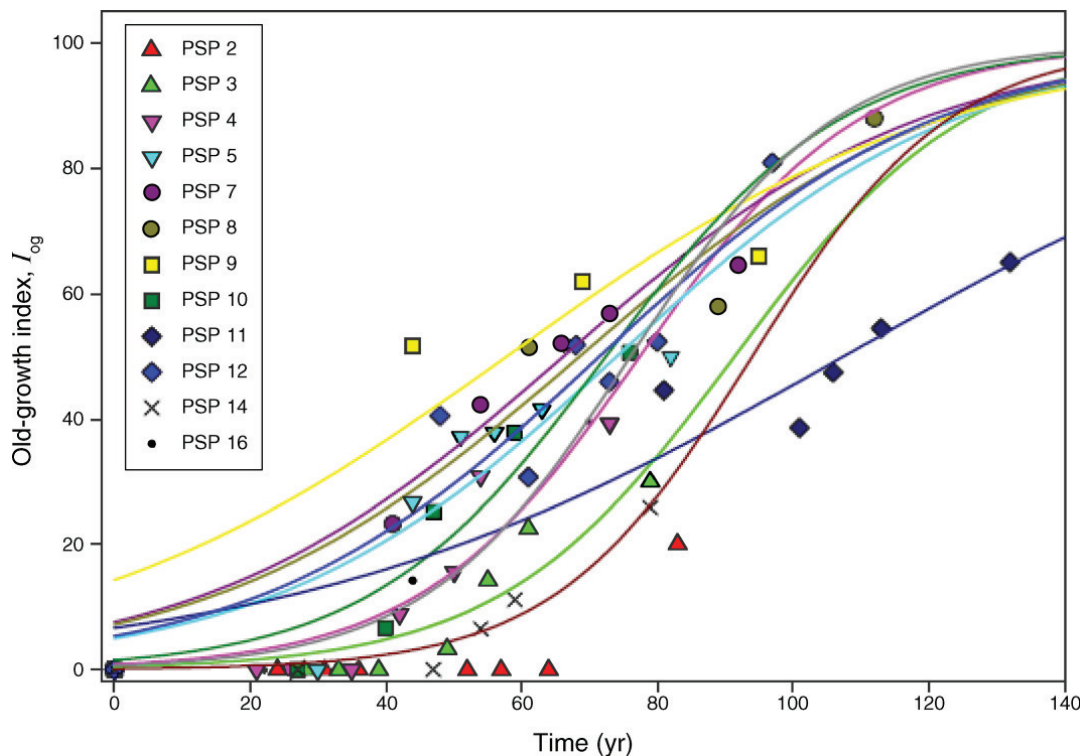


FIG. 2. Developmental trajectory of forest structure as measured by the old-growth index (I_{og}) for permanent sample plots ($n = 12$ plots) and fitted curves ($n = 11$ plots; see Eq. 2). The old-growth index is a unitless composite index based on four measures of live-tree forest structure: standard deviation of dbh for all trees >5 cm dbh; density of Douglas-fir >100 cm dbh (no. trees/ha); mean dbh (cm) for all trees >5 cm dbh; and density of all trees >5 cm dbh (no. trees/ha). The old-growth index is scaled from 0 to 100, with the lower and upper bounds corresponding to published values for regional mean young and old-growth live-tree forest structure variables. See *Methods* for additional detail on the calculation of the old-growth index.

and experienced virtually no net change over the study period.

The final two-dimensional NMS solution (Fig. 7) accounted for 93% of the variance in tree species composition based on the coefficient of determination between the distance in ordination space and distance in the original space for axis 1 ($r^2 = 0.60$) and axis two ($r^2 = 0.33$). The gradient represented by axis 1 was primarily driven by hardwood basal area ($r = -0.86$) and stem density ($r = -0.77$). Axis one was less strongly but negatively related to stem density ($r = -0.43$) and basal area ($r = -0.11$) of shade-tolerant conifers. Stem density and basal area of shade-intolerant conifers showed divergent patterns of association with axis 1: shade-intolerant conifer stem density ($r = -0.36$) and basal area ($r = 0.33$) were weakly correlated with axis 1. Axis 2 was driven primarily by the basal area of shade-tolerant conifers, which was positively related ($r = 0.83$), and stem density of shade-intolerant conifers, which showed a negative relationship ($r = -0.61$). Angiosperm stem density ($r = -0.37$) and basal area ($r = -0.24$) exhibited weak negative relationships with axis 2, while shade-intolerant conifer basal area ($r = 0.41$) and shade-tolerant conifer stem density ($r = 0.45$) showed slightly stronger positive relationships with axis 2.

Forest structure variables exhibited the strongest correlations with axis 2 (Fig. 7). Late-successional forest structure, as measured by I_{og} , was positively correlated with axis 2 ($r = 0.79$). The mean and standard deviation of tree diameter had positive correlations of similar strength with axis 2: $r = 0.75$ and $r = 0.83$, respectively. Density of all trees >5 cm dbh was negatively correlated with both axis 2 ($r = -0.52$) and axis 1 ($r = -0.45$). Density of trees >100 cm dbh was weakly correlated with axis 2 ($r = 0.24$) and not related to axis 1 ($r = 0.01$).

DISCUSSION

High-productivity sites developed toward complex structure more rapidly than low-productivity sites following stand-initiating disturbance in this study, supporting the resilience-productivity hypothesis (Franklin et al. 2002). The proposition that low-productivity ecosystems will be slower to recover from anthropogenic perturbation relative to their high-productivity analogues (Larson and Paine 2007) is also supported, given that the study stands originated from timber harvest.

Tree growth and demographic processes both operate throughout forest structural development. However, the nonuniform increase of the component I_{og} subscores suggests an underlying structure in the contributions of

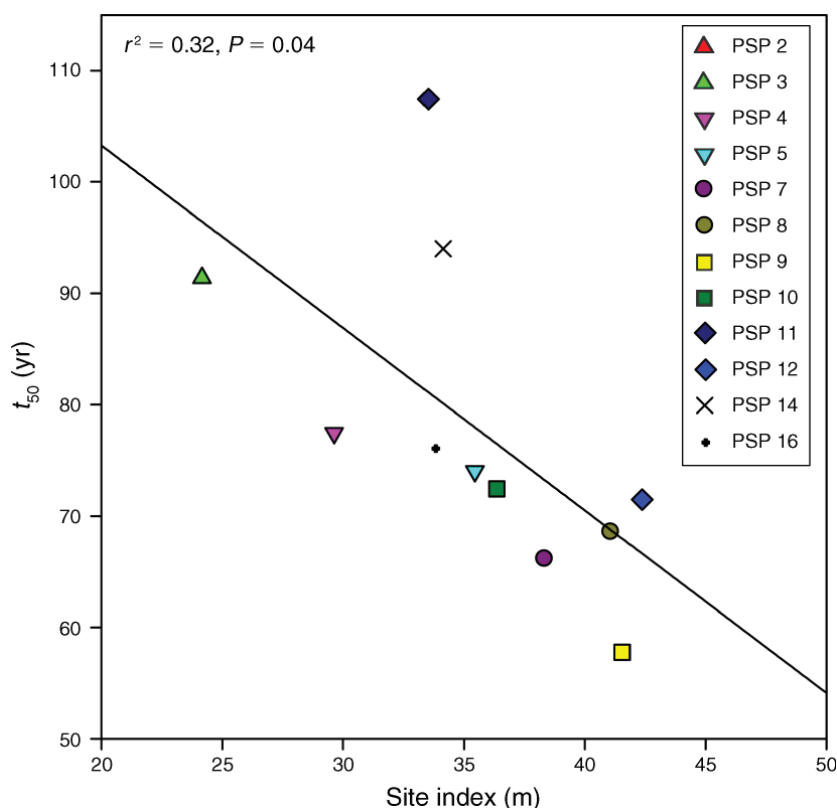


FIG. 3. Regression of time in years to reach $I_{og} = 50$ (t_{50}) on site index, a measure of site productivity. Site index corresponds to the total height (m) of dominant Douglas-fir trees at breast height age of 50 years. Age at breast height and total height of dominant Douglas-fir trees were compared with the empirical age–height curves (King 1966) to determine the site index number for individual study plots. Forest resilience is measured by t_{50} , an estimate of which is provided as a parameter of the fitted logistic growth curve for each study plot (see *Methods*). Lower values of t_{50} correspond to greater resilience.

different processes throughout the structural development sequence: tree growth dominates initially, followed by demographic processes, with the importance of tree growth reemerging later in stand development. The early contribution of the mean tree dbh subscore to the total I_{og} score (Fig. 4) indicates that tree growth initially drives the development of forest structure. As mean tree size increases, competitive tree mortality reduces the density of the initial tree cohort, as reflected by the I_{og} subscore for density of trees >5 cm dbh (Fig. 4). Non-competitive mortality (e.g., due to wind, insects, or pathogens) removes members of the initial cohort during and after the self-thinning phase of stand development (Bible 2001, Lutz and Halpern 2006), providing growing space for shade-tolerant species to recruit (Franklin et al. 2002) and grow (Winter et al. 2002) in the understory, which increases the standard deviation of tree dbh (Fig. 4). The role of competitive and non-competitive tree mortality and shade-tolerant conifer recruitment underscore the importance of demographic processes—tree mortality and recruitment—during intermediate stages of structural development. The late emergence of the trees >100 cm dbh subscore (Fig. 4) illustrates that growth processes are important throughout the developmental sequence, not just in the early stages. Other studies have documented similar trends in

forest structural development. Acker et al. (1998) found that the density of trees >5 cm dbh and mean tree dbh approached old-growth values relatively quickly, while the standard deviation of tree dbh and density of trees >100 cm dbh attained old-growth values later in stand development. Zenner (2004) concluded that structural complexity in old Douglas-fir stands was primarily driven by the standard deviation of tree dbh and density of trees >100 cm dbh.

The strongest successional signal detected in this study was the decline of hardwood tree abundance (Fig. 7). Interestingly, abundance of hardwood species showed little relationship to the development of forest structure. This result contrasts with that of Deal et al. (2004) who concluded that mixed conifer–alder stands are more structurally heterogeneous than pure conifer stands of the same age. It may be that the lack of a relationship between hardwood tree abundance and forest structural development in this present study is due to the relatively low abundance of hardwood trees in the study stands. The lack of a relationship between NMS axis one and forest structural complexity confirms that succession and structural development are not equivalent processes.

Variables representing increasing structural complexity were positively correlated with the second NMS axis,

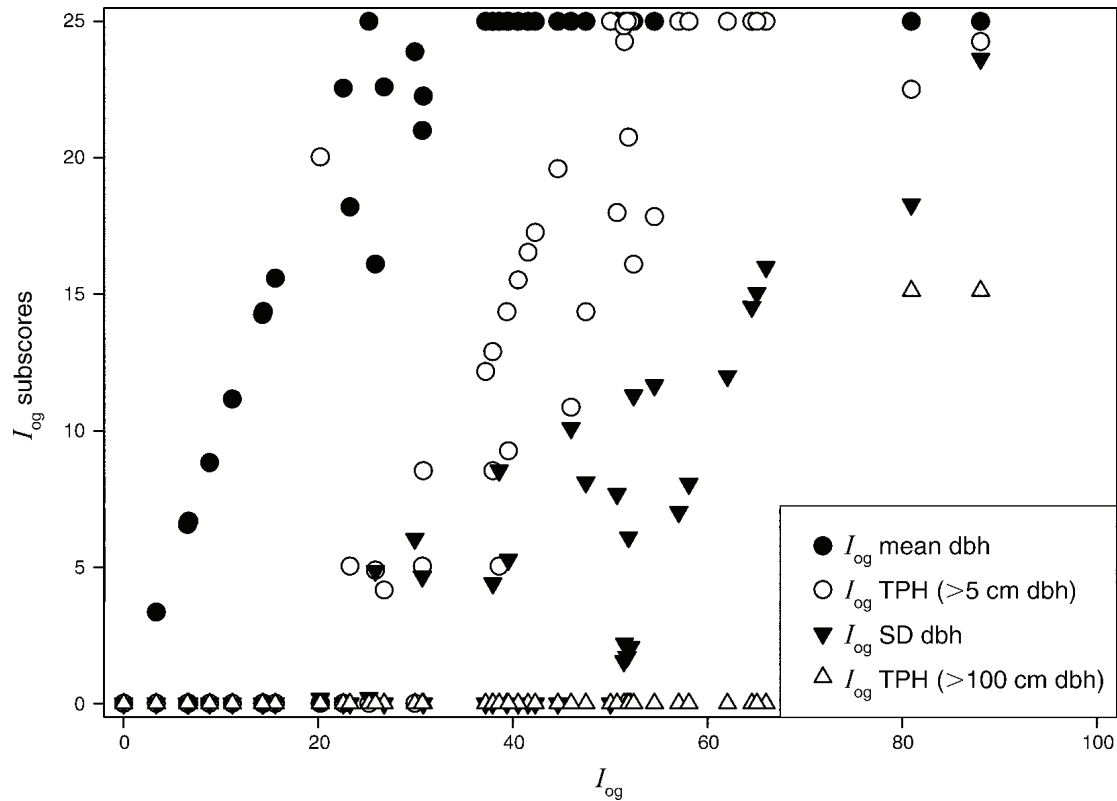


FIG. 4. The individual components of I_{og} each have their period of maximum change during different periods of forest development. The mean dbh subscore is responsible for generating most of the initial nonzero I_{og} values. The subscores for density of trees >5 cm dbh (TPH, number of trees per hectare), standard deviation of dbh, and density (TPH) of trees >100 cm dbh contribute to I_{og} at increasingly later stages of stand structural development, respectively.

the gradient of which was most strongly correlated with basal area of shade-tolerant conifers (Fig. 7). Density of shade-tolerant conifers, however, showed only a moderate positive correlation with axis two. Structural development of Douglas-fir forests occurs concurrent with the increasing abundance of shade-tolerant conifers, primarily western hemlock and western redcedar, as succession advances (Franklin and Hemstrom 1981, Huff 1995, Franklin et al. 2002, Van Pelt and Nadkarni 2004, Keeton and Franklin 2005, Zenner 2005). However, shade-tolerant conifers are not restricted to a late-successional role. They are often present in the stand-initiating cohort (Isaac 1943, Franklin and Hemstrom 1981, Huff 1995, Larson and Franklin 2005, Lutz and Halpern 2006); structurally simple young forest stands can include hundreds of shade-tolerant conifers per hectare (Fig. 6). Structurally complex older forests are distinguished by the presence of large (>40 cm dbh) individual shade-tolerant conifers (Franklin and Spies 1991, Acker et al. 1998), consistent with the positive correlation of forest structure variables with NMS axis two.

The association of the variables contributing to I_{og} , and I_{og} itself with NMS axis two emphasize the necessary distinction between tree species succession and structural development. While succession and

structural development occur contemporaneously, the rates of development between the two are not linearly related and the relationships vary throughout stand development. Here, the loss of three hardwood tree species during succession had little relationship to forest structural development.

Development of structural complexity in Douglas-fir forests requires mortality in the pioneering cohort and recruitment of shade-tolerant conifer species into the lower and middle canopy (Franklin et al. 2002, Zenner 2005). We hypothesize that the observed positive influence of potential site productivity on the rate of forest structural development is in part due to elevated rates of tree population processes—mortality and recruitment—in forests situated on high productivity sites.

Competition-based tree mortality should be positively related to the rate of structural development. The pioneer cohort develops towards a lower number of large individuals during the self-thinning stage of forest structural development (Oliver and Larson 1996, Franklin et al. 2002), creating higher values of I_{og} by both reducing stand density and increasing mean tree diameter (Fig. 4). Elevated rates of competitive mortality on more productive sites have been demonstrated (Yoda et al. 1963, Morris and Myerscough 1991) and

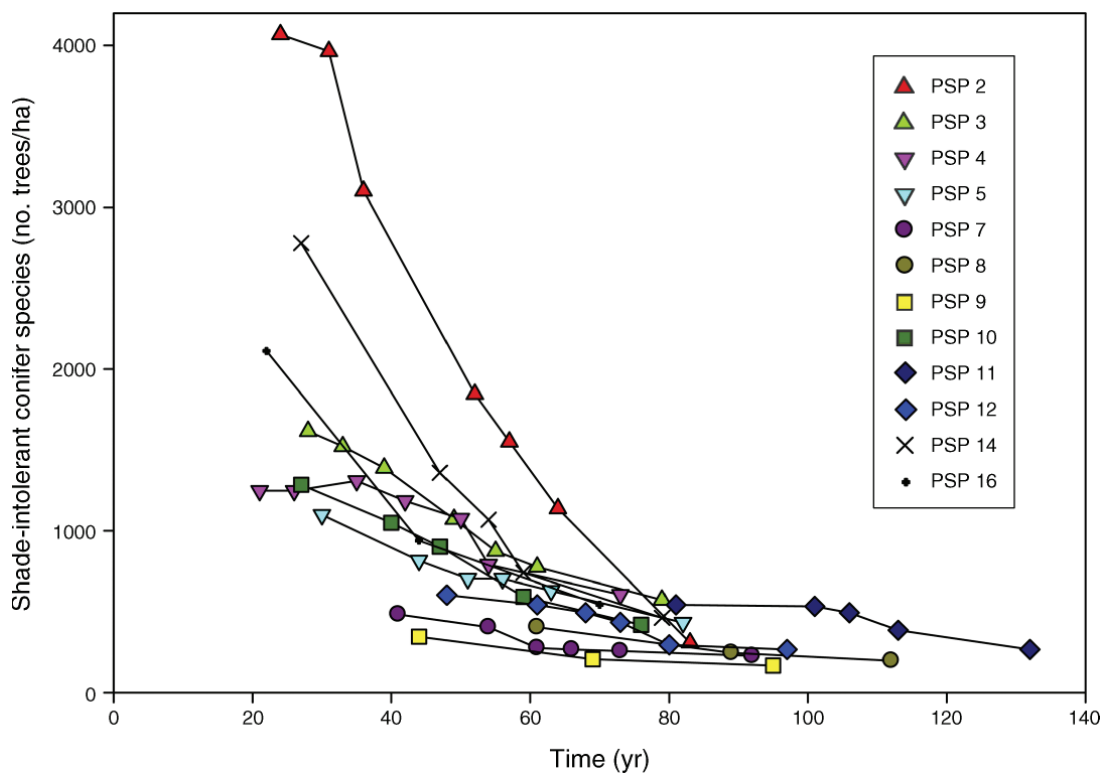


FIG. 5. Changes in density of shade-intolerant conifers (pooled for Douglas-fir, noble fir, Sitka spruce, and western white pine) over time.

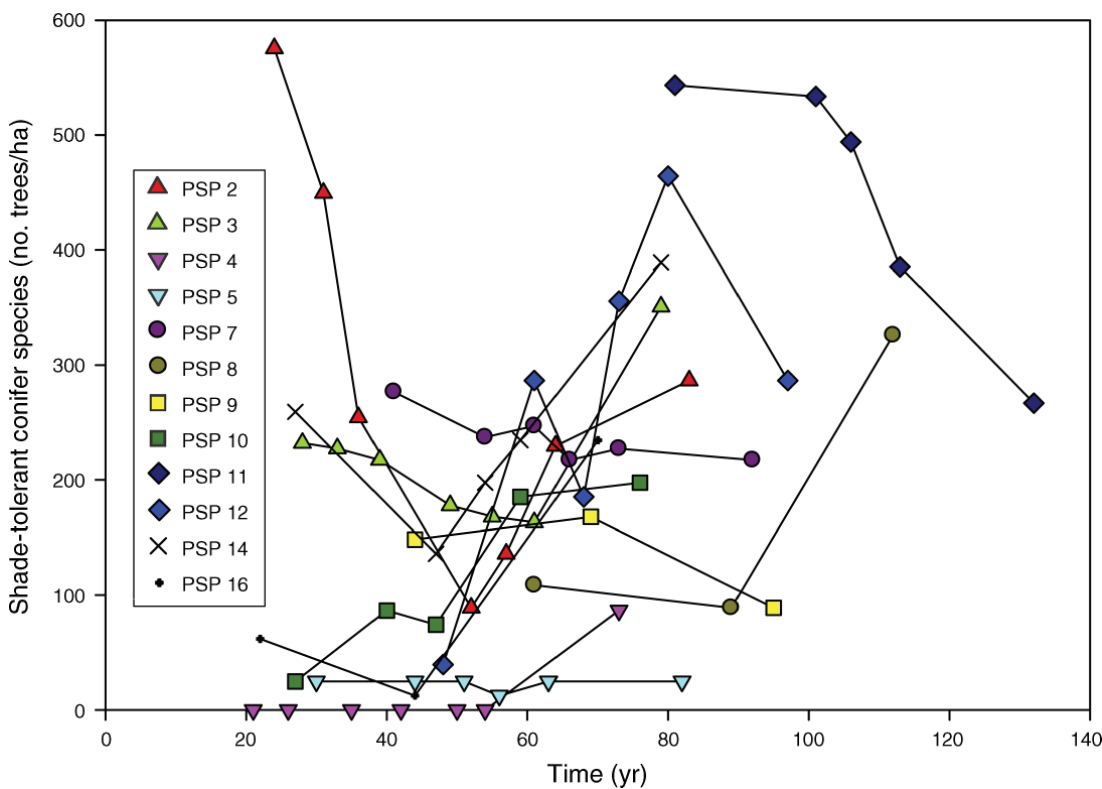


FIG. 6. Changes in density of shade-tolerant conifers (pooled for western hemlock, western redcedar, and Pacific silver fir) over time.

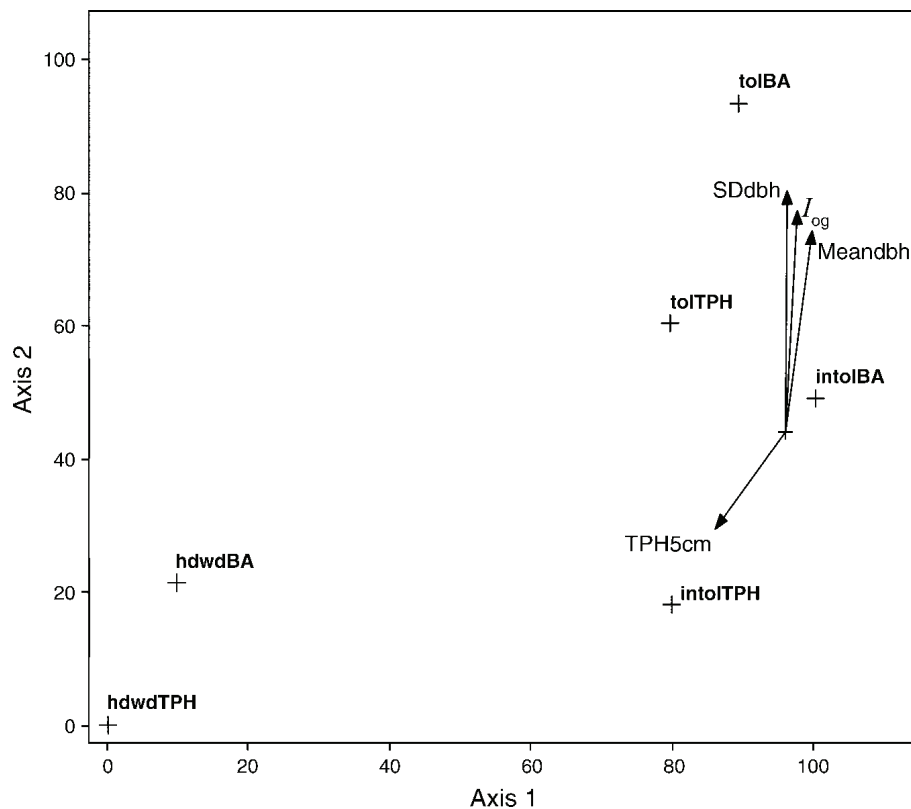


FIG. 7. Nonmetric multidimensional scaling ordination of species composition throughout the study period. Location in ordination space (in boldface type) of hardwood density, in trees per hectare (hdwdTPH), hardwood basal area (hdwdBA), shade-intolerant density (intoI TPH), shade-intolerant basal area (intoI BA), shade-tolerant density (toI TPH), and shade-tolerant basal area (toI BA) is plotted, along with the correlations to the numeric values giving rise to the I_{og} subscores (mean dbh, standard deviation of dbh (SDdbh), TPH for trees >5 cm dbh (TPH5cm); TPH for trees >100 cm dbh not shown) and the calculated composite I_{og} score (TPH, no. trees/ha).

probability of mortality has been shown to be positively related to potential site productivity (Bi 2001, Eid and Tuhus 2001, Shen et al. 2001, Yao et al. 2001). The increased probability of mortality is a consequence of the higher rate of biomass accumulation on higher productivity sites, which is in turn associated with a more quickly decreasing stem density according to the self-thinning relationship (Yoda et al. 1963). Elevated rates of competitive mortality on high-productivity sites decrease the duration of the competitive exclusion phase of stand development, allowing later stages characterized by relatively greater structural complexity to be reached in a shorter period of time.

Episodic noncompetitive tree mortality in the overstory results in growth increases of understory shade-tolerant trees (Winter et al. 2002), which promotes recruitment from the seedling (<5 cm dbh) size class to the tree (>5 cm) size class in canopy gaps (Stewart 1986). Recruitment of shade-tolerant conifers into the tree size class increases within-stand variation in tree diameter (Huff 1995, Zenner 2005), which corresponds with increasing I_{og} scores (Fig. 4). Rates of shade-tolerant tree recruitment should be elevated on high-productivity sites compared to low-productivity sites

given similar rates of non-competitive overstory tree mortality across the productivity gradient; noncompetitive mortality is likely to have variable relationships with productivity depending on species, location, age, and the specific agent of mortality (Franklin et al. 1987). Canopy gaps large enough to influence shade-tolerant tree recruitment develop in a shorter period of time on high-productivity sites because trees attain large size relatively faster on high-productivity sites. This same mechanism is consistent with the observation of Franklin et al. (2002) that spatial heterogeneity due to gap creation develops more rapidly on high-productivity sites.

MANAGEMENT IMPLICATIONS

Currently, a common management objective on many public and some private lands is to develop and maintain structurally complex late-successional forests in landscapes formerly managed for timber production. Active silvicultural treatments, sometimes termed "restoration," are increasingly utilized in young, previously harvested stands to hasten forest structural development (Spies et al. 1991, Carey 2003, Keeton 2006, Davis et al. 2007, Deal 2007).

Site productivity is an important consideration in prioritizing young, previously harvested stands for restoration treatments. Stands on low-productivity sites move through the developmental sequence at a relatively slow pace. Therefore, where reserve-wide reestablishment of structurally complex forests is the primary objective (e.g., City of Seattle 2000) but budgets are limiting, low-productivity sites may have the highest priority for restoration treatments. High-productivity sites should respond rapidly to restoration treatments, producing structurally complex forests in the shortest time. Hence, stands on high-productivity sites may take priority for treatment if the primary objective is restoration of complex structure on a portion of a large management area, for example to provide critical habitat for species of concern.

In many temperate forest ecosystems only small areas of structurally complex forest remain. As a result, conservation reserves are established in previously harvested forests (for example, Late-successional Reserves in the Northwest Forest Plan [USDA and USDI 1994]) so that the desired conditions and associated biodiversity can be created and sustained, respectively, in the future. Results from this study suggest that preserving more productive areas within forest landscapes will yield complex forest structure in a shorter time period. Thus, the productive capacity of forest lands should be considered when designating new reserves, in addition to reserve spatial arrangement and extent. Within the Douglas-fir region variation in site class (Isaac 1949) will be a useful guide for designating new areas to be managed for development of structurally complex forests. More generally, when conservation reserves are designated to allow for the future development of desired habitats, the capacity of the site to develop the desired condition should be included in reserve design and site selection.

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