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1	Effects of sample preparation on methylmercury concentrations in Arctic organisms
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3	Braaten, Hans Fredrik Veiteberg A, B
4	Harman, Christopher ^A
5	Øverjordet, Ida Beathe ^C
6	Larssen, Thorjørn ^A
7	
8	^A Norwegian Institute for Water Research (NIVA), Gaustadalléen 21, N-0349 Oslo, Norway
9	^B University of Oslo (UiO), Department of Chemistry, Sem Sælands vei 26, N-0371 Oslo, Norway
10	^C SINTEF Materials and Chemistry, Marine Environmental Technology, N-7465 Trondheim, Norway
11	* Corresponding author: hbr@niva.no
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Abstract

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The biogeochemical cycling of mercury (Hg) in the marine environment is an issue of global concern since consumption of marine fish is a major route of human exposure to the toxic specie methylmercury (MeHg). The most widely utilized and accepted technique for preparing biological tissue samples for the analysis of MeHg involves an alkaline digestion of the sample. Recent studies suggest however, that this technique is inadequate to produce satisfactory recoveries for certain biological samples, including fish, fur, feathers and other "indicator" tissues which contain relatively high levels of MeHg. Thus an improved acidic extraction method has been proven to produce more satisfactory results for a wide range of biological tissues. The present study compares the two methods on real sample material from different organisms of an Arctic marine food chain, and shows how this could lead to misinterpretation of analytical results. Results show significantly (p < 0.05) lower concentrations for alkaline digestion for large parts of the food chain; especially in fish and birds. The mean differences in concentrations found between the two different methods were 28, 31 and 25 % for fish (Polar and Atlantic cod), Little Auk and Kittiwake, respectively. For samples lower in the food chain (i.e. zooplankton and krill) no significant differences were found. This leads to a clear underestimation of the levels of MeHg found higher up in these food chains; the ratio of MeHg to Hg in biological samples; and thus potentially erroneous conclusions drawn from these results concerning the biological cycling of mercury species. We hypothesize that the main reasons for these differences are poor extraction efficiency and/or matrix effects on the ethylation step prior to analysis. This is the first study to examine the effects of these artefacts on real environmental samples covering a complete food chain.

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Key words

Methylmercury, mercury speciation, sample treatment, acid extraction, alkaline digestion,

53 Arctic food chain

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Introduction

Inorganic mercury (Hg) can undergo methylation into the toxic and bioaccumulative specie methylmercury (MeHg; [1]). MeHg is accumulated in the aquatic food chain with potential harmful effects to organisms [2, 3]. For humans [4], the Hg toxicity is primarily linked to intake of Hg through fish consumption due to high levels of MeHg in predatory fish at the top of the food chain [1], [5]. This is of particular importance in the Arctic where concentrations of Hg in marine animals are about 10-12 times higher than in pre-industrial times [6], and due to the high levels of fish consumption amoung indegineous peoples [7]. To be able to understand the biogeochemical cycling of Hg and human exposure to MeHg toxicity in the marine environment, it is important to choose suitable methodology at all steps of the analytical procedure including; sample collection, sample pre-treatment and measurement [8].

To analyse [9] and isolate [8] MeHg from biological matrices, a wide range of methods have been developed. Historically, digestion of biota in a potassium hydroxide – methanol (KOH-MeOH) solution has been widely employed [1, 10, 11], and has been thought to be an efficient method for extraction. Due to the required dilution (i.e. only small volumes of digest possible for analysis; larger volumes reduces ethylation efficiency) this approach is limited to samples containing high concentrations of MeHg, but method detection limits (MDL) as low as 1 ng/g have been reported [1]. A relatively recent study describes optimal conditions for the alkaline digestion; 5 mL of 15-25 % KOH-MeOH are added to 20 mg sample and heated at 55-60 °C for 24 hours [12].

However, more up to date work has elucidated the shortcomings of the alkaline digestion technique, showing it inadequate to produce satisfactory recoveries for certain biological tissues [12], [13]. The suggested reason for inadequate recoveries when applying alkaline digestion is interference from the organic matrix [12], potentially due to presence of keratin filaments or similar substances [13]. For these reasons an acidic extraction method producing adequate recoveries has been suggested [12, 14]. Briefly this method involves the addition of 10 mL 4.3 M nitric acid (HNO₃) to 20 mg of sample, which is heated at 55-60 °C for 24 hours.

The basis for this study is a set of samples from the Norwegian Arctic covering different organisms of the food chain (zooplankton, fish and birds). All samples were analysed for MeHg

utilizing the alkaline digestion method and the acidic extraction technique for sample pre-treatment. Generally, we hypothesize that concentrations of MeHg will be significantly lower when utilizing the alkaline digestion method compared to the acid extraction method. More specifically, that differences in concentrations will be largest for biological material that contain large amounts of fat, e.g. liver samples, due to the increased interference seen from the organic matrix on analysis of these samples.

Materials and methods

Study area and sampling

Organisms of the pelagic food chain were collected from two fjords at the Svalbard Archipelago in the Norwegian Arctic, Kongsfjorden and Liefdefjorden. Samples were collected in 2007 and 2008 as described in detail elsewhere [15, 16, 17]. Specifications related to collection, preparation and transport are also described in the literature [15, 16, 17 and 18].

A total of 63 samples were used in the present study (Table 1), specifically zooplankton and krill samples (*Calanus finnmarchicus, Calanus hyperboreus, Thiemisto libellula, Thiemisto abysorum*; n = 15), fish samples (Polar cod (*Boreogadus saida*) and Atlantic cod (*Gadus morhua*); n = 20) and marine bird samples (Little Auk (*Alle alle*); n = 16, and Kittiwake (*Rissa tridactyla*); n = 15). The zooplankton and krill samples were extracted and digested as whole species, the Polar cod as whole individual fish, the Atlantic cod and the Little Auk as muscle samples and the Kittiwake as both muscle and liver samples (Specifications in Table 1).

All samples were analysed freeze dried.

Sample preparation

All real samples were extracted and analysed utilizing both an alkaline digestion method and an acid extraction method. Several acids have previously been tested for the purpose of MeHg extraction from different matrices, including HNO₃ [12, 14, 19], sulphuric acid [12], and hydrochloric acid [19, 20]. Our chosen method of acid extraction is based on [12]. In short, samples (minimum 0.03 g) were weighed out, added 10 mL 30 % HNO₃ and heated at 60 °C overnight (approx. 15 hours). Before analysis the extraction solution was added 10 mL DI water. 0.050 mL extraction solution was

neutralized with 0.050 mL 15 % KOH and ethylated before purge/trap and gas chromatography – cold vapor atomic fluorescence spectrometry (GC-CVAFS) analysis and detection as described below.

Two different alkaline digestion methods were tested on reference materials. Alkaline digestion method 1 is based on traditional methods of MeHg biota digestion; described in previous literature [1, 8, 19]. Optimal conditions for this method are described in [12]; samples (minimum 0.02 g) are weighed out, added 5 mL 20 % KOH-MeOH and heated at 55 °C for 24 hours. 50 μ L digestion solutions were analysed without dilution, ethylated before purge/trap and GC-CVAFS analysis and detection.

The other alkaline digestion method that was tested (alkaline digestion method 2) is similar to method 1 and described in [13]. In short, samples are weighed out (minimum 0.02 g), added 1 mL 25 % MeOH-KOH and heated at 65 °C for 3-4 hours. After heating samples are diluted to 2.5 mL with MeOH before analysing a maximum of 0.030 mL of digestion solution. Samples were then ethylated before purge/trap and GC-CVAFS analysis and detection.

The two alkaline digestion methods 1 and 2 were compared by analysing various certified reference materials (CRM; n = 4 parallels; DORM-3 fish protein, TORT-2 lobster hepatopancreas and SRM-2976 mussel tissue). Additionally, the CRMs were also extracted utilizing the acid extraction method for comparison.

Of the alkaline digestion techniques, only method 1 was used when digesting and analysing real sample material.

MeHg analysis

The analysis method for MeHg is based on USEPA Method 1630 [21] for determining MeHg in water by distillation, aqueous ethylation, purge and trap, and CVAFS. Automated systems were used for analysis (Brooks Rand Labs MERX automated systems with Model III AFS Detector). The automated MERX setup is standardised for MeHg analysis, including run duration, heating duration, cooling duration and purge duration of all samples corresponding to 5 minutes, 9.9 seconds, 3 minutes and 6 minutes, respectively. All calibrations standards (covering 0.5 - 1000 pg) were within a recovery of 100 ± 15 %, and produces a relative standard deviation (RSD) of < 8 % for the calibration

coefficients (leading to a r^2 for the calibration curve > 0.9999). Average calibration blanks (< 0.5 pg) and standard deviation of calibration blanks (< 1.0 pg), were also satisfactory.

For every batch of MeHg analysis (n = 30 individual samples) quality assurance and quality control (QA/QC) measures included method blanks (n = 4), sample duplicates (n = 3), matrix spikes (n = 3) and CRMs (n = 6). The certified MeHg concentrations of the CRMs used were 0.355 ± 0.056 mg/kg (\pm uncertainty), 0.152 ± 0.013 mg/kg and 28.09 ± 0.31 µg/kg for DORM-3, TORT-2 and SRM-2976, respectively. Samples that were analysed in duplicates were also used for matrix spike samples. Samples chosen for matrix spikes were added 1000 pg (low concentration samples; 1.0 - 100 ng/g; 0.1 mL of 10.0 ng/mL methylmercury hydroxide; MeHgOH) or 10 000 pg (higher concentration samples; 1.0 - 1000 ng/g; 1.0 mL of 10.0 ng/mL MeHgOH) depending on the concentration of the biological sample. Real sample data were not corrected for CRM recoveries.

All analyses were performed at NIVA's laboratory in Oslo, Norway. MDLs are indicated in the Results as appropriate.

Statistical analysis

To test for differences between the results obtained by the alkaline digestion method and the acid extraction method, Student's t-tests were used. All t-tests shown in this study were two-tailed and homogeneity of variance was tested by F-tests. All statistical analysis and calculations were done in JMP 9.0 (SAS) with a significance level $\alpha = 0.05$, unless otherwise mentioned.

Results

Quality assurance and quality control

Comparison of QA/QC for the two alkaline digestion methods 1 and 2 and the acid extraction method shows little variation. Concentrations of MeHg in blank digestions were 1.6 ± 0.8 pg/mL (mean \pm 1 standard deviation), 1.6 ± 1.4 pg/mL and 1.3 ± 0.7 pg/mL for the acid extraction, alkaline digestion 1 and alkaline digestion 2, respectively (Table 2). This translates to detection limits (DL) of 1 pg/mL or better (3 standard deviations of blank concentrations). The actual limit of detection (LOD) and limit of quantification (LOQ) will vary depending on the weight of sample available for analysis.

For sample weights (0.02-0.1 g) and sample treatment methods included in this study, the LOD is in the range of 0.2-1.0 ng/g (3 standard deviations) and LOQ in the range of 0.3-3.0 ng/g (10 standard deviations). There were no significant differences in blank concentrations between the two alkaline digestion methods 1 and 2 (p=0.21), or between either of the two alkaline digestion methods 1 (p=0.81) or 2 (p=0.09) and the acid extraction method (t-test on difference of mean blank concentrations). The results showing the efficiency of the extraction and digestion methods on CRMs are presented in Table 2. Recovery of MeHg for all extractions and digestions were found to be satisfactory (75 – 125 %; Table 2) and no significant difference was seen individually between the three sample treatment methods (p>0.05). The recovery for the three CRMs, DORM-3, TORT-2 and SRM-2976, were within the expected concentration ranges; 0.299-0.411 mg/kg, 139-165 mg/kg and 27.78-28.40 µg/kg, respectively, for all samples (Table 2).

Matrix spikes and sample duplicates values were similar between the acid extraction and alkaline digestion 1 methods. Spike recoveries were 95.9 ± 11.6 % (acid extraction; n = 8) and 90.6 ± 10.8 % (alkaline digestion 1; n = 8) for the two methods. On average, the relative per cent difference (RPD) between duplicate samples was 11.7 ± 8.9 % (n = 8) and 7.6 ± 7.0 % (n = 8) for the acid extraction (range 1.4 - 26.6 %) and alkaline digestion 1 (range 0.4 - 18.4 %), respectively. No significant difference was seen between the digestion and extraction, neither for spike recoveries (p = 0.37) or sample duplicate RPD (p = 0.33).

Comparison of biological samples data

Concentrations of MeHg in the biological samples ranged from 1.4 - 1050 ng/g. Increasing concentrations were found following the food chain from plankton (including zooplankton and krill; 1.4 - 8.1 ng/g) through fish (Polar and Atlantic cod; 11.4 - 109 ng/g) and Little Auk (140 - 649 ng/g) to Kittiwake (46.7 - 1050 ng/g; individual specie concentration data discussed in Ruus et al. [18]). The mean concentrations of MeHg in plankton (zooplankton and krill), fish (Polar and Atlantic cod), Little Auk and Kittiwake was 3.8 ± 2.1 ng/g, 45.2 ± 27.2 ng/g, 323.5 ± 152.1 ng/g and 545.1 ± 416.8 ng/g, respectively (acid extraction data). See Ruus et al., [18] for a detailed discussion of the biological implications of MeHg concentrations and biomagnification through this Arctic marine food chain.

From Figure 1 it is clear how the acid extraction method generally produced higher concentrations than the alkaline digestion method (levels of MeHg shown as concentration obtained by the acid extraction method divided by concentrations obtained by the alkaline digestion method). In Figure 1 data is grouped in the following manner; 1) plankton (covering both zooplankton and krill); 2) fish (Polar and Atlantic cod); 3) Little Auk and; 4) Kittiwake. The individual sample differences (Table 1) show that the concentrations obtained by the alkaline digestion method averaged 1.7, 28.4, 31.0 and 24.9 % lower than the concentrations obtained by the acid extraction method for the four groups 1) - 4), respectively. No significant difference between the two methods was seen when comparing mean concentrations of the four species groups individually (p > 0.05 for all groups), but concentrations obtained by the alkaline digestion method were lowest in 56 out of 66 samples (Figure 1).

When comparing individual concentration values, the data obtained from the two sample treatments are significantly different for both all data treated together (p < 0.0001) and when grouped as fish (p < 0.0001), Little Auk (p < 0.0001) and Kittiwake (p < 0.001) individually (t-test on difference of paired samples, significance level $\alpha = 0.05$, Wilcoxon Signed Rank). For plankton no significant difference was found (p = 0.81).

No significant relationship was found between concentrations of MeHg and average difference between the two different sample treatment methods when all samples were included ($r^2 = 0.05$). The same applies to analysing the relationship on individual species groups ($r^2 < 0.10$) and when MeHg concentrations are studied on a logarithmic scale ($r^2 < 0.10$; data not shown).

Discussion

Quality assurance and quality control

Comparison of the two alkaline digestion methods and the acid extraction method indicates similar performance for all three methods regarding QA/QC data, including method blanks, sample duplicates and matrix spikes. The similar performance related to MeHg concentrations in all quality data indicate that both methods theoretically should be suitable for determining concentrations of MeHg in biological samples.

Based on the blank concentrations (Table 2) and weight of biological samples (0.02 - 0.1 g) we derive LODs of 0.2 - 1.0 ng/g for both the alkaline digestion method 1 and the acid extraction method (3 standard deviations of method blanks). This is low enough for most biological samples from Arctic food chains and similar to what other studies are documenting utilizing the same sample treatment methods (alkaline digestion; [1, 8, 12, 19], acid extraction; [12]) and analytical techniques. However, six of our samples are below the LOQ (S2, S4, S5, S7, S8 and S9; Table 1). These are all plankton samples and are still included in the study (Figure 1). They do not affect the interpretation and main conclusion of our findings.

Our study also shows that there is no significant difference in the recovery efficiency between the different extraction methods for the three CRMs that we used; DORM-3, TORT-2 and SRM-2976. Compared to the biological data in the present manuscript, we conclude that our three CRMs represent appropriate concentration ranges for MeHg bioaccumulation studies of Arctic food chains. Thus in studies where the alkaline digestion is used, acceptable CRM recoveries may provide false confidence in actual sample extraction recoveries. Differences between CRM recoveries and actual sample recoveries are likely to be largest where the two biological matrixes are most different, for example feather samples versus fish muscle tissue (DORM-3). In the absence of CRMs for all matrices this means that care must be taken when interpreting results from one type of sample when the CRM is formulated from a different type of sample. An alternative approach may be to investigate the use of a surrogate internal standard, more usually applied in classical organic chemistry analysis. Both ethyl and propyl mercury have been used for this purpose, especially when applying GC-MS analysis [22, 23], although this approach is also not without its challenges.

While DORM-3 is widely used in studies of MeHg in biological matrices, use of the other two CRMs we utilized is less prevalent. However, when analysing low concentrations of MeHg in biological material, the relatively low concentrations of the SRM-2976 means it is a good option for quality control.

Concentrations of MeHg found in our biological samples from Svalbard are of similar levels that are previously shown for Arctic food chains [6], and representative of marine food chains. No samples were below MDL and concentrations span 3 orders of magnitude; 1.4 – 1050 ng/g. This gives a good basis for studying the different efficiency of sample treatment methods; i.e. alkaline digestion versus acid extractions.

In the present study our data shows how concentrations of MeHg obtained from the alkaline digestion method are always lower than concentrations obtained by the acid extraction method as long as concentrations exceed 10 ng/g (specific concentrations not shown here; see Ruus et al. [18] for details). For low level samples (< 10 ng/g) the difference is not significant (p > 0.05). This could be due to either the nature of the low concentration samples (plankton and krill) or the fact that concentrations are close to MDL levels in general. The same pattern of small or no difference between the two sample treatment methods can be seen also for samples of fish and Kittiwake with low concentrations.

For samples exceeding 10 ng/g the concentrations obtained by the alkaline digestion are significantly lower than for the acid extraction concentrations. There is relatively little variation in RPD between concentrations for the different species groups of the food chain; fish (including both Polar and Atlantic cod), Little Auk and Kittiwake; 28.4, 31.0 and 24.9 %, respectively. Previous studies suggest that lower recoveries are seen for the alkaline digestion method due to interference from the organic matrix [12, 13]. Hintelmann and Nguyen [12] suggests that the alkaline digestion method does not completely decompose the organic matrix. An incomplete decomposition leads to intact functional groups in the solution which later interefere with the analytical procedure.

However, if such a matrix effect exists, this effect should potentially increase down the food chain as the ratio of animal protein to intereferring substance (i.e. organic matrix) potentially decreases. In this study we do not observere such a pattern. In fact no relationship was found between concentrations of MeHg and average difference between the two different sample treatment methods $(r^2 < 0.1)$ when all samples were included and for individual species groups). This indicates that the relatively poor recovery of MeHg when using the alkaline digestion method compared to the acid extraction is relatively unaffected by concentration magnitude. I.e. the alkaline digestion sample pre-

treatment method always produces unsatisfactory and significantly lower results compared to the acid extraction method. However, it can be seen that the largest absolute differences are found when concentrations of MeHg are the highest.

Interestingly, there is no or little difference seen in extraction efficiency when comparing the Atlantic cod results from this study (Figure 1), whereas for Polar Cod there were significant differences for concentrations > 20 ng/g (n = 12; p < 0.001). The Polar Cod (n = 16) samples were composed from whole individuals, whereas the Atlantic cod samples (n = 4) were taken from muscle tissue only. This appears to support the idea that the alkaline extraction method is not efficient for matrices which are more complex than fish muscle.

Comparing the two sample preparation techniques for Kittiwake samples reveals that 5 samples are responsible for the largest absolute differences seen between the two techniques. These 5 samples (marked with asterisk in Figure 1) are all liver samples and therefore have a relatively high fat content. Again, this corresponds with the idea that the efficiency of the alkaline digestion technique will decrease with increasing "other organic matrix". Although not relevant for liver samples, one such suggested interference is keratin filaments. These may be present in large amounts in certain types of biological samples, such as hair, fur and feathers [13]. Whilst these matrices are not considered in the present study this is an important point as these types of sample are widely analysed for MeHg due to the desirability of non-destructive sampling. In some cases this interference or poor extraction may be sufficient as to result in MeHg not being detected.

Although comparison of QA/QC for the alkaline digestion method 1 and the acid extraction method shows little variation, we do acknowledge that artefact methylation during the extraction procedure could possibly explain the differences seen for our real samples. Artefact methylation is reviewed in Leermakers et al. [8], where it is stated that previous assessments of the potential to generate MeHg from inorganic Hg during sample preparation are mainly related to distillation-based methods. Our acid extraction method is based on Hintelmann and Nguyen, [12], which does not point to any problems of artefact methylation. Neither is it mentioned as a problem by other studies utilizing this method [13, 14]. For sediment and soil samples both the positive (artefact formation of MeHg) and negative bias (incomplete leaching, and/or decomposition of MeHg) were investigated for a

similar acid extraction method in Liang et al. [24]. In Liang et al. [24], the conclusion is that while the distillation process shows artefact formation of MeHg when concentrations of Hg reaches 2000 ng/g, the nitric acid extraction is independent of Hg concentrations. This confirms the discussion in Liang et al. [8], and while we cannot exclude artefact methylation completely, we conclude that the acid extraction technique is a good option for MeHg concentration ranges found in Arctic marine food chains.

As most studies examining MeHg in aquatic food chains (freshwater and marine), today and historically, use powdered fish or other similar CRMs to imply extraction efficiency, the effect highlighted in the present study go unnoticed. Where total Hg (TotHg) and MeHg are both measured in the same sample and concentrations of the organic form are found to be lower, then it is assumed that this is a real difference. Thus it is widely reported that whilst almost all Hg in fish (muscle) is present in the organic form [1], in many other biological samples the ratio is lower [5]. This may however, just be an artefact from the extraction method chosen and thus may easily lead to a significant underestimation overall of MeHg concentrations through the food chain. We suggest therefore that levels of MeHg may in general be under-reported in the literature, where the alkaline digestion method has been applied. This in turn may result in erroneous conclusions about the fate of mercury species in biological food chains, and our understanding of the bio-geochemical cycling of MeHg in the environment in general.

Conclusions

Results from the present study show significantly different MeHg concentrations in biological samples, depending on the extraction method used. Although the two methods produce comparable QA/QC results (blanks, duplicates, spikes and available CRMs) the concentrations found in most biological matrices, especially fish and bird liver, are significantly different. This led to an underestimation of between 24 and 31 % of the MeHg concentrations in the studied food chain. Such differences may have consequences for our understanding of the bioaccumulation of Hg species, their speciation in biota, and thus their biogeochemical cycling in general. Based on present results we conclude that care must be taken when choosing the sample treatment method for analysis of MeHg in

biological samples, and that interpretation of results from alkaline digestions should be carried out with caution.

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360 Literature

- 361 [1] BLOOM, N. S. On the chemical form of mercury in edible fish and marine invertebrate tissue.
- 362 Canadian Journal of Fisheries and Aquatic Sciences, 49, 1010-1017, 1992.
- 363 [2] WHO. Environmental Health criteria 101, Methyl Mercury. International Programme on Chemical
- 364 Safety. WHO, Geneva, Switzerland, 1991.
- 365 [3] SCHEULHAMMER, A. M., MEYER, M. W., SANDHEINRICH, M. B. & MURRAY, M. W.
- Effects of environmental methylmercury on the health of wild birds, mammals, and fish.
- 367 *Ambio*, 36, 12-18, 2007.
- 368 [4] MERGLER, D., ANDERSON, H. A., CHAN, L. H. M., MAHAFFEY, K. R., MURRAY, M.,
- 369 SAKAMOTO, M. & STERN, A. H. Methylmercury exposure and health effects in humans: A
- worldwide concern. *Ambio*, 36, 2007.
- 371 [5] UNEP. Global Mercury Assessment. UNEP Chemicals Branch, Geneva, Switzerland, 2002.
- 372 [6] UNEP. United Nations Environment Programme. Global Mercury Assessment 2013: Sources,
- Emissions, Releases and Environmental Transport. UNEP Chemicals Branch, Geneva,
- 374 Switzerland, 2013.
- 375 [7] DEUTCH, B., DYERBERG, J., PEDERSEN, H. S., ASCHLUND, E. & HANSEN, J. C.
- 376 Traditional and modern Greenlandic food Dietary composition, nutrients and contaminants.
- *Science of the Total Environment*, 384, 106-119, 2007.
- 378 [8] LEERMAKERS, M., BAEYENS, W., QUEVAUVILLER, P. & HORVAT, M. Mercury in
- environmental samples: Speciation, artifacts and validation. Trac-Trends in Analytical
- 380 *Chemistry*, 24, 383-393, 2005.
- 381 [9] SUVARAPU, L. N., SEO, Y. K. & BAEK, S. O. Speciation and determination of mercury by
- various analytical techniques. *Reviews in Analytical Chemistry*, 32, 225-245, 2005.
- 383 [10] LIANG, L., HORVAT, M., CERNICHIARI, E., GELEIN, B. & BALOGH, S. Simple solvent
- 384 extraction technique for elimination of matrix interferences in the determination of
- methylmercury in environmental and biological samples by ethylation gas chromatography
- cold vapor atomic fluorescence spectrometry. *Talanta*, 43, 1883-1888, 1996.

- 387 [11] BLOOM, N. Determination of picogram levels of methylmercury by aqueous phase ethylation,
- followed by cryogenic gas-chromatography with cold vapor atomic fluorescence detection.
- *Canadian Journal of Fisheries and Aquatic Sciences*, 46, 1131-1140, 1989.
- 390 [12] HINTELMANN, H. & NGUYEN, H. Extraction of methylmercury from tissue and plant samples
- by acid leaching. *Analytical and Bioanalytical Chemistry*, 381, 360-365, 2005.
- 392 [13] BROOKS RAND LABS. Determining methylmercury concentrations in mammals and birds
- 393 utilizing nondestructive sample collection techniques. Report. Brooks Rand Labs, Seattle,
- 394 WA, USA, 2012.
- 395 [14] HAMMERSCHMIDT, C. R. & FITZGERALD, W. F. Methylmercury in arctic Alaskan
- mosquitoes: implications for impact of atmospheric mercury depletion events. *Environmental*
- 397 *Chemistry*, 5, 127-130, 2008.
- 398 [15] HALLANGER, I. G., RUUS, A., HERZKE, D., WARNER, N. A., EVENSET, A., HEIMSTAD,
- E. S., GABRIELSEN, G. W. & BORGA, K. Influence of season, location and feeding strategy
- on bioaccumulation of halogenated organic contaminants in Arctic marine zooplankton.
- 401 Environmental Toxicology and Chemistry, 30, 77-87, 2011a.
- 402 [16] HALLANGER, I. G., RUUS, A., WARNER, N. A., HERZKE, D., EVENSET, A., SCHOYEN,
- 403 M., GABRIELSEN, G. W. & BORGA, K. Differences between Arctic and Atlantic fjord
- 404 systems on bioaccumulation of persistent organic pollutants in zooplankton from Svalbard.
- 405 *Science of the Total Environment*, 409, 2783-2795, 2011b.
- 406 [17] HALLANGER, I. G., WARNER, N. A., RUUS, A., EVENSET, A., CHRISTENSEN, G.,
- HERZKE, D., GABRIELSEN, G. W. & BORGA, K. Seasonality in contaminant
- 408 accumulation in Arctic marine pelagic food webs using trophic magnification factor as a
- measure of bioaccumulation. Environmental Toxicology and Chemistry, 30, 1026-1035,
- 410 2011c.
- 411 [18] RUUS, A., ØVERJORDET, I. B., BRAATEN, H. F. V., EVENSET, A., GABRIELSEN, G.,
- BORGÅ, K. Methylmercury biomagnification in an Arctic food web a sidelong glance
- across species, seasons and locations. *In preparation*.

414	[19] HORVAT, M., LIANG, L., AZEMARD, S., MANDIE, V., VILLENEUVE, J. P. & COQUERY,
415	M. Certification of total mercury and methylmercury concentrations in mussel homogenate
416	(Mytilus edulis) reference material, IAEA-142. Fresenius Journal of Analytical Chemistry,
417	358, 411-418, 1997.
418	[20] FERNANDEZ-MARTINEZ, R. & RUCANDIO, I. A simplified method for determination of
419	organic mercury in soils. Analytical Methods, 5, 4131-4137, 2013.
420	[21] USEPA (United States Environmental Protection Agency). Method 1630: Methylmercury
421	in Water by Distillation, Aqueous Ethylation, Purge and Trap, and Cold Vapor Atomic
422	Fluorescence Spectrometry (ed. W. A. Telliard. United States Environmental
423	Protection Agency - Office of Water), 1998.
424	[22] GIBICAR, D., LOGAR, M., HORVAT, N., MARN-PERNAT, A., PONIKVAR, R. & HORVAT,
425	M. Simultaneous determination of trace levels of ethylmercury and methylmercury in
426	biological samples and vaccines using sodium tetra(n-propyl)borate as derivatizing agent.
427	Analytical and Bioanalytical Chemistry, 388, 329-340, 2007.
428	[23] YAN, D., YANG, L. M. & WANG, Q. Q. Alternative thermodiffusion interface for simultaneous
429	speciation of m organic and inorganic lead and mercury species by capillary GC-ICPMS using
430	tri-n-propyl-lead chloride as an internal standard. Analytical Chemistry, 80, 6104-6109, 2008.
431	[24] LIANG, L., HORVAT, M., FENG, X. B., SHANG, L. H., LIL, H. & PANG, P. Re-evaluation of
432	distillation and comparison with HNO3 leaching/solvent extraction for isolation of
433	methylmercury compounds from sediment/soil samples. Applied Organometallic Chemistry,
434	18, 264-270, 2004.
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437	Tables
438	
439	Table 1. Individual sample specifications.
440	<i>Table 2.</i> QC/QA for the sample treatment techniques.
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Table 1. Specifications (species group, species name and analytical sample specification) with levels of MeHg shown as concentration obtained by the acid extraction method divided by concentrations obtained by the alkaline digestion method. Individual concentration differences are shown as relative percent difference (RPD). ID numbers refer to sample ID used throughout the paper. Sample IDs in bold are below limit of quantification (LOQ, n = 6).

ID	1 Species name		Analytical sample specification	[Acid extraction] : [Alkaline digestion]	RPD (%)
S1	Plankton	Calanus finmarchicus	Whole individuals	1.0	3.4
S2	Plankton	Calanus hyperboreus	Whole individuals	0.6	45.1
S3	Plankton	Calanus sp	Whole individuals	11.4	167.7
S4	Plankton	Calanus sp	Whole individuals	0.7	34.7
S5	Plankton	Calanus sp	Whole individuals	0.5	65.6
S6	Plankton	Calanus sp	Whole individuals	1.0	1.1
S7	Plankton	Calanus sp	Whole individuals	1.2	16.4
S8	Plankton	Calanus hyperboreus	Whole individuals	0.8	21.8
S9	Plankton	Calanus hyperboreus	Whole individuals	0.5	69.7
S10	Plankton	Themisto libellula	Whole individuals	0.9	11.0
S11	Plankton	Themisto libellula	Whole individuals	1.4	35.7
S12	Plankton	Themisto abysorum	Whole individuals	1.1	8.0
S13	Plankton	Calanus sp	Whole individuals	1.5	41.6
S14	Plankton	Calanus sp	Whole individuals	1.1	5.3
S15	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.0	0.4
S16	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.0	2.2
S17	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.4	31.4
S18	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.2	20.9
S19	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.1	9.5
S20	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.1	10.2
S21	Bird	Kittiwake (Rissa tridactyla)	Muscle	0.7	33.2
S22	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.1	13.3
S23	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.7	52.0
S24	Bird	Kittiwake (Rissa tridactyla)	Muscle	1.5	37.5
S25	Bird	Kittiwake (Rissa tridactyla)	Liver	1.7	49.3
S26	Bird	Kittiwake (Rissa tridactyla)	Liver	1.2	22.0
S27	Bird	Kittiwake (Rissa tridactyla)	Liver	1.8	59.0
S28	Bird	Kittiwake (Rissa tridactyla)	Liver	1.7	50.1
S29	Bird	Kittiwake (Rissa tridactyla)	Liver	1.7	49.4
S30	Bird	Little Auk (Alle alle)	Muscle	1.3	28.0
S31	Bird	Little Auk (Alle alle)	Muscle	1.3	27.3
S32	Bird	Little Auk (Alle alle)	Muscle	1.3	27.7
S33	Bird	Little Auk (Alle alle)	Muscle	1.2	18.7
S34	Bird	Little Auk (Alle alle)	Muscle	1.4	34.4
S35	Bird	Little Auk (Alle alle)	Muscle	1.1	13.0
S36	Bird	Little Auk (Alle alle)	Muscle	1.4	30.7
S37	Bird	Little Auk (Alle alle)	Muscle	1.3	26.6
S38	Bird	Little Auk (Alle alle)	Muscle	1.3	27.0
S39	Bird	Little Auk (Alle alle)	Muscle	1.2	21.6
S40	Bird	Little Auk (Alle alle)	Muscle	1.6	44.0
S41	Bird	Little Auk (Alle alle)	Muscle	1.5	41.0
S42	Bird	Little Auk (Alle alle)	Muscle	1.6	48.7
S43	Bird	Little Auk (Alle alle)	Muscle	1.5	42.9
S44	Bird	Little Auk (Alle alle)	Muscle	1.4	33.3
S45	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.3	25.8
S46	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.6	45.6
S47	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.5	38.3
S48	Fish	Polar cod (Boreogadus saida)	Whole individuals	2.5	87.1
S49	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.7	50.2
S50	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.5	38.6
S51	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.6	46.5
S52	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.3	26.8
S53	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.1	13.1
S54	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.4	31.0
S55	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.4	32.3
S56	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.2	15.2
S57	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.2	15.5
S58	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.0	1.5
S59	Fish	Polar cod (Boreogadus saida)	Whole individuals	1.4	35.9
S60	Fish	Atlantic cod (Gadus morhua)	Muscle	1.3	22.7
S61	Fish	Atlantic cod (Gadus morhua)	Muscle	1.1	5.3
S62	Fish	Atlantic cod (Gadus morhua)	Muscle	1.2	16.3
S63	Fish	Atlantic cod (Gadus morhua)	Muscle	0.9	5.6

Table 2. Certified reference material QC/QA for the comparison of the three different sample treatment methods; acid extraction (AE), alkaline digestion 1 (AD 1) and alkaline digestion 2 (AD 2).

	Cample	n	MeHg concentration		% recovery
Sample	Sample treatment method		Average	Standard deviation	(average ± 1 standard deviation)
CRM	AE	15	0.325 mg/kg	0.027 mg/kg	91.6 ± 7.7
(DORM-3)	AD 1	8	0.313 mg/kg	0.024 mg/kg	88.2 ± 6.7
(DOKNI-3)	AD 2	4	0.298 mg/kg	0.024 mg/kg	83.9 ± 6.7
CRM	AE	15	0.158 mg/kg	0.011 mg/kg	104.2 ± 7.3
(TORT-2)	AD 1	8	0.146 mg/kg	0.009 mg/kg	95.8 ± 5.6
(10K1-2)	AD 2	4	0.148 mg/kg	0.010 mg/kg	97.6 ± 6.9
CDM	AE	15	27.25 μg/kg	0.89 µg/kg	97.0 ± 3.2
CRM (SRM-2976)	AD 1	8	24.62 μg/kg	1.38 µg/kg	87.6 ± 4.9
(SKW1-2970)	AD 2	4	23.62 µg/kg	1.66 µg/kg	84.1 ± 5.9

Figures

Figure 1. MeHg ratios in the biological samples.



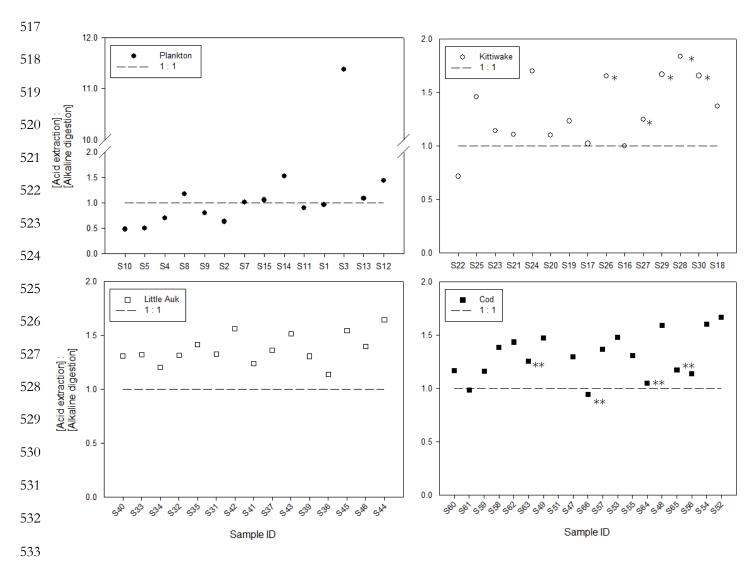


Figure 1. Levels of MeHg in the biological samples as concentration obtained by the acid extraction divided by concentrations obtained by the alkaline digestion. Figure shows plankton samples (top left), Kittiwake samples (top right), Little Auk samples (bottom left) and Polar and Atlantic cod samples (bottom right). The dotted horizontal lines represent the 1:1 relationship between the concentrations obtained by the two sample treatment techniques. Samples are sorted by increasing concentrations of MeHg obtained by the acid extraction method from left to right. Kittiwake liver samples (n = 5) are indicated by a single asterisk and Atlantic cod samples (n = 4) are indicated by two asterisks.