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1 **Evaluation of stand level hybrid PipeQual-model with permanent sample plot data of**
2 **Norway spruce**

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22 Abstract

23 The evaluation of process-based models (PBM) includes ascertaining their ability to produce
24 results consistent with forest growth in the past. In this study, we parameterized and
25 evaluated the hybrid model, PipeQual, with datasets containing traditional mensuration
26 variables collected from permanent sample plots (PSP) of even-aged Norway spruce (*Picea*
27 *abies* (L.) Karst) stands in Finland. To initialize the model in middle of stand development
28 and reproduce observed changes in Norway spruce crown structure, the built-in empirical
29 relationships of crown characteristics were made explicitly dependent of light environment.
30 After these modifications, the model accuracy at the whole dataset level was high, slope
31 values of linear regressions between the observations and simulations ranging from 0.77 to
32 0.99 depending on the variable. The average bias in stand dominant height ranged between -
33 0.72 – 0.07 m, -0.68 – 0.57 cm in stand mean diameter, -2.62 – 1.92 m² in stand basal area
34 and 20 - 29 m³ in stand total stem volume. Stand dynamics after thinning also followed
35 reasonable closely the observed patterns. Accurate predictions illustrate the potential of
36 model for predicting forest stand growth and forest management effects in changing
37 environmental conditions.

38 Keywords

39 *Picea abies*, growth simulation model, validation, forest management effect, growth response

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45 **Introduction**

46 The importance of forests as resources of raw material is growing due to increasing pressure
47 to reduce the use of fossil fuels and fossil based materials. At the same time, the
48 environmental change driven by climate change challenges traditional silvicultural practices,
49 creating an increasing demand for tools capable of predicting forest responses to both climate
50 and management. Empirical growth models are generally thought to be of limited usefulness
51 under changing conditions since their representation of environmental effects is, at its best,
52 implicit and derived from past data. (Monserud 2003; Cuddington et al. 2013; Lonsdale et al.
53 2015). Process-based forest growth models (PBM) are based on a mechanistic description of
54 processes which in principle allows for projections into the future once the driving variables
55 and process parameters have been determined. In practice, however, no model is fully
56 mechanistic and some degree of model calibration against data on predicted variables is
57 always required (Korzhuikun et al. 1996; Mäkelä et al. 2000; Robinson and Ek 2003;
58 Monserud 2003; Fontes et al. 2010 Cuddington et al. 2013). This adds an empirical
59 dimension to PBMs, making them “hybrid” to a lesser or greater degree (Korzhuikun et al.
60 1996; Mäkelä et al. 2000), and emphasizes the need for thorough evaluation of such models
61 against available data as a prerequisite for any future predictions (e.g. Vanclay and
62 Skoovsgard 1997; Bokalo et al. 2013). The evaluation should quantitatively assess how
63 useful the model is for the specific purpose it has been constructed.

64 Evaluation of process-based or hybrid models is far from simple. There are two main
65 challenges in the evaluation. Firstly, many of these models do not incorporate variables
66 describing explicit tree and stand structure which could be evaluated against empirical stand
67 data. Secondly, strict testing demands proper datasets which are scarce. Yield tables and
68 inventory data have been used in PBM evaluation (e.g., Lonsdale et al. 2015; Mäkelä et al.
69 2016). However, permanent sample plots (PSPs) provide more rigorous data for the

70 evaluation. To date, a few studies exist that have evaluated process-based or hybrid forest
71 growth models against data from PSPs (Robinson and Ek 2003; Raulier et al. 2003; Zhou et
72 al. 2005; Pinjuv et al. 2006; Fontes et al. 2006; Minunno et al. 2010). From the viewpoint of
73 evaluating PBMs and hybrid models with PSP data, a specific challenge is that most of the
74 PSPs have been established after juvenile stage in stands of pole or mature stage. The PBMs
75 usually contain a large number of state variables to be initialized, but only a subset of these is
76 available in the PSP data. The PBMs are therefore usually initialized at the seedling or
77 sapling stage where forest management has not yet affected tree or stand structure (e.g.,
78 Pérez-Cruzado et al. 2011) and required state variables are easier to attain. If stand
79 management prior to the first PSP measurement has been recorded, the initial state can be
80 estimated through simulation from stand establishment to the first measurement, possibly
81 combined with some calibration (e.g. Kantola et al. 2007). However, this method is
82 problematic for model applications as the management history of forest stands is usually
83 unknown and actual management pathways vary considerably.

84 In the hybrid stand growth model PipeQual (Mäkelä et al. 1997; Mäkelä and Mäkinen 2003;
85 Kantola et al. 2007), the initial data requirements are low because the model uses structural
86 constraints that connect standard forestry variables with each other and functional biomasses.
87 In addition to these constraints, we hypothesized here that connecting the tree and crown
88 structure more tightly to stand light conditions would allow us to account for the adaptation
89 of tree structure occurring from sapling to subsequent developmental stages and solve this
90 way the initialization problem in the middle of stand development.

91 The objective of this study was to test the PipeQual model for spruce against PSP data in
92 southern and central Finland, with special focus on requirements outlined below. In
93 particular, we aimed at modifying the model to make it applicable from any initial state
94 measured in the field, regardless of stand age. In case the results deviate from the

95 measurements, our aim was to interpret the causes of this in terms of model assumptions.
96 Finally, we interpreted the results from the perspective of evaluating PBMs against PSP data
97 in general.

98 In our evaluation, we thus concentrated on the tree and stand characteristics easily
99 measurable in the field and which are of importance in forestry practice. To pass the
100 evaluation successfully, PipeQual should fulfill the following requirements:

101 REQ1: The model must be initializable at any time during the rotation.

102 REQ2: The model must be unbiased with respect to annual volume increment.

103 REQ3: Simulated growth responses to forestry operations, e.g. thinning, must be qualitatively
104 and quantitatively reasonable .

105 REQ4: Tree form and growth allocation to different compartments must behave logically
106 over time and in response to forestry operations.

107

108 **Material and methods**

109 **Description of the model**

110 The PipeQual model is based on ecological theory and describes stand and tree growth as a
111 result of carbon acquisition and allocation (Mäkelä 1997, 2002). The model consists of four
112 modules, STAND, TREE, WHORL and BRANCH, through which tree metabolism, tree
113 structure and stand dynamics are interconnected in the framework of carbon balance at an
114 annual time resolution (Fig. 1 Structure of model, Supplementary material). The Norway
115 spruce version of the model is described in detail in Kantola et al. (2007), and subsequent
116 modifications reported in Niinimäki et al. (2012, 2013) and Mäkelä et al. (2016). Here, we

117 describe the key characteristics of the model, as well as outline some further modifications
118 made in this study.

119 The stand is described as a distribution of tree size classes. Each class is represented by its
120 mean tree and stocking density. Annual photosynthetic production is first computed for the
121 whole stand and then allocated to trees using a modified Lambert-Beer equation (Duursma
122 and Mäkelä 2007). This is input to the TREE module where the growth of trees is derived
123 from the carbon balance of the mean tree of the size class. The mean tree acquires carbon,
124 respire, and loses biomass through turnover. Growth is allocated to foliage, branches, stem,
125 coarse roots and fine roots to maintain a regular structure derived from the pipe model
126 (Shinozaki et al. 1964), profile theory (Chiba et al. 1988) and fractal crown allometry
127 (Mäkelä and Sievänen 1992; Duursma et al. 2010). These regularities allow for tracking the
128 development of dimensional variables in addition to the biomass variables.

129 Climate and site impacts enter the model through (1) tissue-specific rates of carbon fluxes
130 and (2) carbon allocation coefficients. Through modularity of model structure, different
131 submodels may be chosen to describe these processes. Here, we use the approach introduced
132 by Mäkelä et al. (2016) who proposed to incorporate these effects in PipeQual under current
133 climate by means of effective temperature sum (ETS) and site type (as defined by Cajander
134 1949). Potential gross primary production depends on ETS because temperature, radiation
135 (photosynthetically active radiation, PAR) and vapour pressure deficit (VPD) correlate with
136 each other while soil water availability only plays a minor role in Norway spruce in Finland
137 under current climate (Härkönen et al. 2010; Minunno et al. 2016). Growth respiration in the
138 model is proportional to net production while maintenance respiration depends on the
139 biomass of different compartments, air temperature and nitrogen concentration of tissue
140 (through site type). Tissue life span is also related to climate and nitrogen content. Mäkelä et
141 al. (2016) demonstrated that the consequent climate and soil driven geographic trends of

142 forest carbon balance components yielded accurate estimates of the productivity of Norway
 143 spruce stands in Fennoscandian conditions.

144

145 **Modifications to model structure**

146 In the original PipeQual model (Mäkelä et al. 1997, 2000), stand density effects on tree
 147 structure were described on the basis of crown coverage which mediated both crown rise and
 148 mortality. No other density impacts on tree and stand structure were included. In order to
 149 improve the description of tree interactions in the model, here foliage density and crown
 150 width were made responsive to the competitive position of the tree. We made these explicitly
 151 dependent on the light availability to the trees instead of crown coverage.

152 The proportion of photosynthetically active light reaching height x in the canopy is calculated
 153 in the model as

$$f(x) = \exp \left[- \int_H^x \sum_i k_{i,\text{eff}} l_i(y) dy \right] \quad (1)$$

154

155 where H is canopy height, l_i is the vertical leaf area distribution of tree class i , $k_{i,\text{eff}}$ is the
 156 effective extinction coefficient of tree class i (Duursma and Mäkelä 2007). The mean light
 157 environment experienced by tree class i is characterised as

$$f_{Mi} = \frac{1}{2} [f(H_i) + f(H_{Ci})] \quad (2)$$

158

159 where H_i and H_{Ci} are the height and crown base height of tree class i , respectively.

160 Using these definitions, crown rise was made dependent on the light conditions at the base of
 161 the crown. Firstly, we assumed the rise of the crown to follow the height growth as suggested
 162 by Valentine and Mäkelä (2005):

$$\frac{dH_C}{dt} = s_C \frac{dH}{dt} \quad (3)$$

163

164 and secondly, we made the coefficient s_C dependent on the proportion of light reaching the
 165 base of the crown of the tree class in a ramp-like manner as follows:

$$s_C = 1 - \frac{1}{1 + e^{-a(f(H_C) - f_0)}} \quad (4)$$

166

167 Here, a and f_0 are parameters (Table S1). In other words, crown rise occurs if the light level
 168 below the crown goes below a threshold, then rapidly accelerates to match height growth as
 169 the light levels fall (Figure S1).

170 In addition, plasticity of the crowns was assumed in two parameters. The foliage density
 171 parameter ξ was allowed to decline in trees in very poor light as follows:

$$\xi = \min \left\{ \xi_0, \xi_0 \frac{f_M}{f_1} \right\} \quad (5)$$

172 where ξ_0 is the value of this parameter in good and moderate light conditions and f_1 is the
 173 relative light level below which foliage density declines (Table S1).

174 Secondly, the growth of branch length was assumed to be regulated by the stand crown
 175 coverage A_{tot} ; in sparser stands crowns were assumed to grow wider than in denser stands:

$$\gamma_b = \min \left\{ \gamma_{b0}, \gamma_{b0} \frac{A_{\text{tot},0}}{A_{\text{tot}}} \right\} \quad (6)$$

176

177 where γ_b is the steady-state crown radius to crown length ratio, γ_{b0} is that ratio in a sparse
178 stand, and $A_{\text{tot},0}$ is a parameter (Table S1). When changes occur in A_{tot} the model changes γ_b
179 gradually, tracking the effect of increased carbon demand of the accelerated growth of
180 branches. This new formulation replaced the dependence of γ_b on tree age assumed by
181 Niinimäki et al. (2012).

182

183 **Description of permanent sample plots**

184 The *Harkas* dataset of Natural Resources Institute Finland used in model evaluation consisted
185 of 19 stands containing altogether 126 PSPs. These PSPs included 30800 sample trees of
186 which 20576 Norway spruce were used in this evaluation (Table 1, Mäkinen and Isomäki
187 2004a, b). Temperature sum (ETS) of sites ranged from 1033 to 1275 d.d. and forest site type
188 of 18 stands was fertile (*Oxalis-Myrtillus*) and one stand was classified as medium fertile
189 (*Myrtillus*) site type according to the Finnish classification system (Cajander 1949). The
190 stands were even-aged, almost pure Norway spruce stands. Measurements were taken 3 - 8
191 times, the longest measurement period being 37 years. In each measurement, tree species,
192 stem diameter at 1.3 m (D1.3) and possible damage were measured in each tree. Tree height,
193 height of the crown base and stem diameter at 6 m were measured, on average, from 54
194 sample trees in each plot. However, crown base heights had not been measured in the first
195 measurement in any of the PSPs. Stand age at the first measurement ranged from 26 to 57
196 yrs, stand density from 935 to 3335 ha⁻¹, and dominant height (average height of the hundred

197 thickest trees in the stand) from 9.8 to 24.3 m. In the last measurement, stand dominant height
198 ranged from 29 to 34.5 m.

199 Each stand included plots of different intensity of thinning from below – unthinned, light,
200 moderate, and heavy thinning. The number of plots per stand ranged mostly from 4 to 12, i.e.,
201 the four treatments were replicated in a randomized block design. Treatment intensity was
202 defined on the basis of the basal area of the thinned plot relative to that of the unthinned
203 control plot (*c.f.* Mäkinen and Isomäki 2004a). This dataset was utilized for re-calibrating the
204 model after the modifications made to it in this study (see Modifications to model structure).

205 Another smaller PSP dataset (*Syst*, Table 1, Mäkinen et al. 2006) contained altogether 1565
206 trees on two experiments representing the fertile *Oxalis-Myrtillus* forest site type. In this
207 dataset, stand age at the first measurement ranged from 38 to 41 yrs, stand density from 2780
208 to 3500 ha⁻¹, and dominant height from 15.8 to 16.8 m. The experiments were measured five
209 times within a time frame of 17 and 21 years. This dataset was not utilized in model
210 calibration and could thus be regarded as an independent data set.

211 **Initialisation of the model**

212 Because of the structural constraints in PipeQual, the initial state of each tree can be
213 computed from tree height, breast height diameter, crown base height and tree age. The initial
214 values of tree height, D1.3, and tree age were constructed for the size classes of the simulated
215 stands from the first measurement of each PSP. The crown base heights needed for the
216 determination of the crown ratio were not measured in the first measurements of PSPs and
217 were thus determined by an empirical model (see below, eqs. 10 and 11). The diameter
218 distribution of trees in the first measurement of each PSP was formed with two centimeter
219 class width and the mean tree height, crown ratio etc. in each class were computed (SAS 9.4,

220 SAS Institute Inc. 2015). This approach produced 10-20 size classes per plot which were used
221 to describe the stand structure in PipeQual initialization.

222 Tree height and crown ratio are used for deriving the initial foliage and fine root mass and
223 sapwood areas of the woody components in the TREE module (Mäkelä et al. 1997). The
224 WHORL module (Kantola et al. 2007) is thereafter used for initialising the vertical
225 distribution of variables across whorls, including whorl-mean branch length, branch and stem
226 sapwood area and foliage mass. Initial tree age is needed for setting the initial number of
227 whorls. This information is combined with breast height stem diameter for initialising the
228 stem and branch heartwood both in TREE and WHORL modules. Lengths and diameters of
229 individual branches are generated using empirical, stochastic functions for the number of
230 branches in each whorl and a disaggregation of mean basal area and length to individual
231 branches.

232 We predicted initial crown base heights by using the empirical model of Petersson (1997) in
233 its original form. We derived model parameter estimates from stand variables obtained from
234 the second, third etc. measurement of PSPs. The crown base height was measured in the field
235 as the height of the lowest living branch above which no more than one dead whorl exists.
236 Petersson's model utilizes both the tree level variables, like tree height, tree diameter and
237 H/D ratio, and stand level variables, like stand density, basal area, volume and age, in the
238 estimation of the crown base height:

$$239 \ln(y_{ilk}) = b_0 + b_1 \times \ln(x_{ilk}^1) + \dots b_n \times \ln(x_{ilk}^n) + s_k + p_{lk} + \ln(\varepsilon_{ilk}) \quad (7)$$

240 where y_{ilk} is the height of the crown base of tree i in plot l of stand k , b_j and x^j ($j =$
241 $1, \dots n$) are the fixed-effect coefficients and independent variables, respectively, and s_k
242 (stand) and p_{lk} (plot) account for the hierarchical data structure (Table 2). Instead of reducing
243 random variation in stands by taking class means (which were described in terms of size class

244 mean trees), the obtained crown base height estimates of individual trees were smoothed in
245 each stand separately over tree height (h) with the Gompertz function (Fig. 2, Huang et al.
246 1992):

$$247 \quad y = Ae^{[-e^{(\beta-\kappa h)}]} \quad (8)$$

248 where y is the estimated height of the crown base, and A , β and κ are parameters.

249 **Parameterisation**

250 In this study, we estimated values for the parameters included in the new functions
251 introduced above (see Modifications to model structure). The parameters were estimated
252 through hand-tuning of educated initial guesses, with the *Harkas* data as a qualitative
253 reference (Supplementary material). Other model parameters have been reported in previous
254 studies (Kantola et al. 2007; Niinimäki et al. 2012; Mäkelä et al. 2016).

255 **Model evaluation**

256 We simulated the development of each stand from the initial condition, i.e. from the age of
257 the first measurement of each PSP, until the stand reached the age of 100 years. In the
258 simulations, we applied a thinning routine where the same number of trees was removed as in
259 the field in each thinning. The share of removed trees per size class was based on the
260 proportion of stand basal area contained in each size class.

261 Test diagnostics were calculated for both *Harkas* and *Syst* datasets. These included absolute
262 model bias (AMB), relative model bias (RMB), and modeling efficiency (EF) to obtain an
263 overall understanding of model behavior (Table 3, Pinjuv et al. 2006). In addition, we
264 evaluated the accuracy of model predictions by plotting the simulated values against the
265 measured values and fitting a linear regression between the observations and simulations. The
266 slope, significance, and coefficient of determination (R^2) of the regressions were used for

267 determining the accuracy of model predictions (Table 3, Fig. 3). Trends in residuals were
268 inspected in order to observe over or underestimation of the model.

269

270 **Results**

271 Model accuracy at the whole dataset level was acceptable, suggesting that the simulations
272 were able to meet our requirement of initialization at an arbitrary stand age (REQ1). Slope
273 values between simulated and observed values ranged from 0.84 to 0.99 in different variables
274 (Fig. 3). Slopes differed significantly from 1 ($p < 0.001$) in all variables indicating bias in
275 model predictions. However, AMB of stand mean diameter and dominant height were low
276 (Table 3), indicating high overall accuracy. RMB was highest in stand volume ca. 8%.

277 Stand mean stem diameter was slightly overestimated in stands with dbh < 20 cm, while in
278 stands with larger mean diameter it was generally underestimated (Fig. 3a). Stand dominant
279 height was predicted accurately in most of the stands throughout the observed range (Fig. 3b).
280 Simulated stand density followed the observed (Fig. 3c) as should be expected due to the
281 applied thinning routine where the same number of trees was removed as in the actual
282 thinnings. However, this also indicates the success of the applied mortality function (Reineke
283 1933). In only a few unthinned or lightly thinned stands, the model predicted a slightly higher
284 mortality rate than the measured one in the PSPs. The simulated diameter distributions of
285 stands were, however, narrower and more peaked than those observed (Fig. 4). Both stand
286 basal area and volume were slightly underestimated (Fig. 3d and e), the AMB being 1.9 m^2
287 ha^{-1} and $29 \text{ m}^3 \text{ ha}^{-1}$, respectively (Table 3). Stands with the lowest volume were slightly
288 overestimated while stands with high stem volume were generally underestimated and
289 prediction error increased with increasing stand volume.

290 The annual volume increment (volume production) residuals were in average under $1 \text{ m}^3 \text{ yr}^{-1}$
291 (Fig. 5a) which we interpret to be small enough in order for the model to pass REQ2. There
292 was a trend from overestimation in the heavily treated sites to underestimation in the
293 unthinned stands. However, total annual increment also includes drain (harvests and
294 mortality). Separation of these shows that the model tended to underestimate net increment
295 (rate of change of standing volume) and overestimate drain (Fig. 5b and c) but no treatment
296 effect was observed in these terms.

297 The results illustrated in Fig. 5 indicate that the model was able to meet REQ3, at least partly,
298 since the observed trend in residuals between treatments was reasonably small. In a closer
299 look, the simulated stand dynamics generally followed closely the observed patterns both in
300 the unthinned and thinned plots (Fig. 6). However, in some stands (e.g. in Hauho, Fig. 6)
301 stand basal area was clearly overestimated in the unthinned and lightly thinned plots. A
302 similar pattern could be seen in stand volume.

303 For testing the REQ4, we analyzed the simulated height to diameter ratio (H/D), crown base
304 heights and stand leaf area index (LAI). Both simulated H/D and crown base height showed a
305 moderate correspondence with the observations (Fig. 7a and 8a). The development of
306 simulated H/D over stand age followed a logical pattern, responding to thinning as expected
307 although not quite as strongly as observed (Fig. 7b). This deviation was especially clear in
308 heavily treated plots. The height to crown base was generally overestimated at the
309 initialization, the error being largest in the trees with low crown base heights (Fig. 8). The
310 simulated rise of the crown base and thinning response seemed generally to follow closely the
311 observed (Fig. 8b) and no treatment effect was observed in the residuals (data not shown).

312 The stand leaf area index (LAI) estimate was not included in the PSP dataset. Thus, we
313 evaluated the LAI prediction of PipeQual by estimating the foliage mass of simulated trees

314 with empirical biomass models (Marklund 1988; Repola et al. 2009) and converting those
315 values to LAI with the mean specific leaf area value used in PipeQual. PipeQual generally
316 predicted higher stand LAI than these two empirical models (Fig. 9a). In most of the stands,
317 PipeQual and Repola's model predicted similar thinning responses while in Marklund's
318 model the development of LAI after thinning deviated from the other models (Fig. 9b).

319 In the comparison against the smaller independent dataset (*Syst*), the model showed similar
320 accuracy as in the larger *Harkas* PSP dataset (Table 3). The level of stand mean diameter,
321 predicted dominant height and stand basal area did not deviate from the observations ($p >$
322 0.05) while the level of stand volume was underestimated ($p = 0.006$). In stand mean
323 diameter and stand volume, the residuals had a significant trend with stand age ($p = 0.048$
324 and 0.026 , respectively).

325

326 **Discussion**

327 In this study, our main focus was on evaluating the PipeQual model for applications in
328 forestry practice, where data on standard forest mensuration variables are available from
329 stands at an arbitrary measurement age, and the aim is to make predictions about forest
330 growth and yield under different thinning schedules. While most empirical forest growth
331 models have been planned for precisely this type of use, process-based models are prone to
332 face a challenge because of their detailed input requirements and lack of variables relevant to
333 forest mensuration (Cuddington et al. 2013). However, there is a need for models capable of
334 projecting the potential impact of climate change on long-term patterns of forest growth and
335 development while being reasonably accessible to forest managers. Most importantly, such
336 models should be able to represent the effects of climate change on forest productivity, to
337 simulate a variety of forest management options for creation of adaptive management
338 strategies and be relatively straightforward to calibrate (Seely et al. 2015). Hybrid models,

339 like PipeQual here, could be seen as a bridge between traditional statistical forest growth and
340 yield models and overly complex PBMs called for the support in forest management
341 decision-making in changing environmental conditions (Valentine and Mäkelä 2005; Fontes
342 et al. 2010; Cuddington et al. 2013). Climatic envelope models are too simplistic for
343 predicting changes in tree growth at stand level which is the meaningful scale for forest
344 management (Kimmins et al. 2008). Stand level hybrid models are founded in ecological
345 theory and incorporate key ecophysiological processes involved in tree growth rates to deal
346 adequately with the increasing uncertainty of future tree growth and climate change effects
347 on forests (Lo et al. 2010). We see model evaluation as a prerequisite for using any model in
348 forest management decision making or scenario modeling. Model evaluation aims at
349 assessing how well a model performs in the types of application it was planned for (Bokalo et
350 al. 2013).

351

352 In our evaluation, we combined several metrics in order to get a robust idea of the strengths
353 and weaknesses of the model. The low AMB values (Table 3) are indicative of coherence of
354 model predictions over the whole dataset. The high efficiency values indicate that there is
355 little variation in the residual errors, i.e., the predictions had high precision (but not
356 necessarily good accuracy, *c.f.* Bokalo et al 2013). The prediction error in the annual volume
357 increment was at an acceptable level (on average ca. 1 m³), thinning responses were logical
358 and quantitatively reasonable, and the simulated tree form in terms of height to diameter ratio
359 (H/D) showed a fair correspondence with the observations. These results support the
360 robustness of the model and its applicability to assess the effects of forest management
361 activities, while Mäkelä et al. (2016) have shown that PipeQual is also able to reproduce the
362 geographic variation of tree growth within forests of Finland.

363

364 That an arbitrary age and stand structure could be used for model initialization, was primarily
365 due to the inclusion of structural constraints (e.g. the ratio between leaf mass and sapwood
366 area, see Supplementary material) in the model which allowed us to derive all key state
367 variables from a small subset of inputs. However, the new components of the model
368 accounting for acclimation of foliage density and crown width to light conditions further
369 improved model performance especially with regard to the initial state. The model bias was
370 not, however, stable over time since residuals of the examined variables had a trend, with
371 increasing bias over stand age. This trend was the strongest in stand-mean diameter and
372 weakest in stand volume, suggesting that while total stem increment was reasonably well
373 estimated, there may have been problems in the allocation of growth between stem height and
374 diameter.

375

376 Tree form is a measure of the trade-off between accumulating height and accumulating
377 diameter (Robinson and Ek 2003). Trees grown in high stocking have different stem form
378 from those grown in sparse stands (Assman 1970, Fontes et al. 2006). PipeQual produced
379 underestimates in high H/D values (Fig. 7a) due to simulated small trees being shorter than
380 those observed relative to their diameter, whereas simulated large trees were taller than those
381 observed relative to their diameter resulting in overestimation of low H/D values.
382 Nevertheless, the fact that the model explicitly considers stem form as part of the carbon
383 allocation routine is crucial for converting the carbon fluxes into meaningful forest
384 mensuration variables. Most of the PBMs either ignore stem dimensions (e.g., Landsberg and
385 Waring 1997, Pietritsch et al. 2007) or assume a prescribed H/D ratio (Thornley and Cannell
386 2000). In PipeQual, the allocation between height and diameter growth is linked to crown
387 ratio and crown rise according to the profile theory (Chiba et al. 1988), and the model does
388 not make any *a priori* assumptions about the height growth pattern. With no crown rise, more

389 growth is allocated to increasing the size of the crown. This results in a demand of basal area
390 growth through the pipe model assumption. In contrast, when crown recession occurs, more
391 photosynthates are needed for the replacement of the dead branches and the corresponding
392 woody “pipes” as well as for the height increment, whereas crown size and therefore basal
393 area growth remain low (Valentine and Mäkelä 2005).

394

395 Despite the causal structure of this allocation pattern, there are still problems in actually
396 quantifying it. As demonstrated in the present parameter estimation (see also Supplementary
397 material), the resulting H/D ratio is sensitive to the assumptions about crown rise. Similarly,
398 changes in crown shape and the demand of photosynthates to branch growth affect the
399 pattern. The new model features implemented here about the impact of the light climate on
400 foliage density and branch length amplified the effects of crown rise by increasing the
401 photosynthate demand of foliage and branches in high relative to low light availability.
402 Nevertheless, the simulated crown rise was faster than the observed. This could partly be
403 caused by the model of Pettersson (1997) used for predicting initial height of crown base, in
404 some cases, produced initial crown base heights that were higher than the actual heights in
405 the subsequent measurements, although part of the effect evidently came from the model
406 dynamics.

407 Our results suggest that the physiological foundation of PipeQual was sufficient for
408 producing plausible growth responses to thinnings. According to ecological theory and
409 forestry experience, tree populations respond to thinning by increased diameter increment,
410 maintained or reduced height increment and reduced crown recession (Assmann 1970;
411 Vuokila and Väliäho 1980; Pretzsch 2002; however, see Raulier et al. 2003). From the
412 carbon balance modeling perspective this is indicative of changes in carbon allocation after
413 thinning between diameter and height increment on one hand, and between crown rise and

414 crown extension on the other hand. (Mäkelä 1997; Valentine and Mäkelä 2005). Thinning
415 slowed down the simulated crown rise, however, the effect was modest compared with the
416 observations on the heavily treated plots. Heavy thinning yielded a clear halt in the observed
417 crown rise for several years, whereas in the simulations the plateau was clearly shorter. One
418 reason behind this could be that the model overestimates production after heavy thinning
419 (Fig. 5a) since the photosynthesis model does not take into account the clustering of foliage
420 enough (Duursma and Mäkelä 2007).

421

422 The basic carbon balance modeling approach in PipeQual follows the general convention
423 used in many models with different species and conditions (Thornley and Cannell 2000,
424 Fontes et al. 2006). On the other hand, it differs from many models in the way it translates the
425 carbon balance into structural growth. Here, the key components are the assumptions made
426 on conservative structures, such as the pipe model, and on plastic structures, such as crown
427 ratio. Valentine and Mäkelä (2005) presented an explicit derivation of dimensional growth
428 from the carbon flux rates (Bridging Model) under assumptions virtually identical with the
429 CROBAS module of PipeQual. The Bridging Model has been successfully parameterised for
430 stand mean trees using statistical fitting in spacing experiments (Valentine et al. 2012, 2013)
431 and Bayesian calibration in national inventory data (Van Oijen et al. 2013). The PipeQual
432 model has also been previously parameterised for Scots pine (Mäkelä and Mäkinen 2003).
433 These results suggest that the structural assumptions of PipeQual are realistic and generic, but
434 adjustment to actual parameter values will nevertheless be required when moving from one
435 ecosystem to another. Our on-going work in this respect is focusing on deciduous stands, old
436 growth forests and continuous cover management.

437

438 Like in any carbon balance model, the responses to a changing environment are accounted for
439 through the corresponding responses in the carbon flux rates, including the relative allocation
440 of carbon to below-ground components (Mäkelä et al. 2016). The relatively simple
441 description of the metabolic processes compared with many PBMs (Medlyn et al. 2011) and
442 the modular structure of PipeQual ensure its flexible application to different conditions. For
443 example, the module for deriving potential photosynthetic production from effective
444 temperature sum can easily be replaced by a module responsive to a number of environmental
445 drivers (Minunno et al. 2016) that may be of importance under climate change. A more
446 challenging question is that of nitrogen availability in relation to photosynthetic potential
447 (e.g. Mäkipää et al. 2015). As Cuddington et al. (2013) point out climate change may alter the
448 processes which operate at different scales than decision-making and are still critical for
449 model predictions e.g. global atmospheric CO₂ concentration. In addition, there remains
450 uncertainty about how an ecological process will interact with new global change conditions.
451 This challenges the scale and scope of models developed for answering forest management
452 questions. It is evident that continuous evaluation of models will be required as data
453 gradually becomes available from changing environments.

454

455 **Conclusions**

456 Process-based simulation model predictions should be evaluated before their use as a tool for
457 depicting the growth and development of forest stands in a changing environment. In this
458 evaluation, we showed that the hybrid model, PipeQual, was able to describe forest
459 management effects on Norway spruce in terms of traditional stand mensuration variables at
460 an acceptable level of error. Low AMB and RMB values along with sensible thinning
461 responses demonstrate that PipeQual was reliable enough for tasks usually performed by
462 empirical forest models. The structural constraints coupled with underlying processes form a

463 foundation of the model which ensures applicability in different conditions. The new
464 functions connecting tree crown characteristics explicitly to stand light conditions further
465 improved this flexibility. This approach enabled more stringent evaluation of the model with
466 PSP data but it will also enhance the applicability of the model to real life situations where
467 the prediction of stand development from random initial conditions is of interest. The
468 approach provides both forecasts and insights into the underlying processes that drive
469 changes in forest growth, thus helping us answer forest management related questions in
470 scenario modeling of future climate conditions.

471

472 **Acknowledgements**

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474

475 **References**

- 476 Assmann, E. 1970. The principles of forest yield study. Pergamon Press, Oxford, U.K.
- 477 Bokalo, M., Stadt, K.J., Comeau, P.G. and Titus, S.J. 2013. The validation of the mixedwood
478 growth model (MGM) for use in forest management decision making. *Forests* 4:1-27.
- 479 Cajander, A.K. 1949. Forest types and their significance. *Acta For. Fenn.* 56.
- 480 Chiba, Y., Fujimori, T. and Kiyono, Y. 1988. Another interpretation of the profile diagram
481 and its availability with consideration of the growth process of forest trees. *J. Jap. For. Soc.*
482 70:245-254.
- 483 Cuddington, K., Fortin, M.-J., Gerber, L. R., Hastings, A., Liebhold, A., O'Connor, M. and
484 Ray, C. 2013. Process-based models are required to manage ecological systems in a changing
485 world. *Ecos.* 4:20.
- 486 Duursma, R. and Mäkelä, A. 2007. Summary models for light interception and light-use
487 efficiency of non-homogeneous canopies. *Tree Physiol.* 27:859-870.
- 488 Duursma, R.A., Mäkelä, A., Reid, D.E.B., Jokela, E.J., Porté, A., Roberts, S.D. 2010. Self-
489 shading affects allometric scaling in trees. *Funct. Ecol.* 24:723-730.
- 490 Fontes, L., Landsberg, J. Tomé, J., Tomé, M., Pacheco, C.A., Soares, P. and Araujo, C. 2006.
491 Calibration and testing of a generalized process-based model for use in Portuguese eucalyptus
492 plantations. *Can. J. For. Res.* 36:3209-3221.

- 493 Fontes, L., J.-D. Bontemps, H. Bugmann, M. Van Oijen, C. Gracia, K. Kramer, M. Lindner,
494 T. Rötzer and J. P. Skovsgaard 2010. Models for supporting forest management in a
495 changing environment. *For. Sys.* 19:8-29.
- 496 Härkönen S., Pulkkinen M., Duursma R.A., Mäkelä A. 2010. Estimating annual GPP, NPP
497 and stem growth in Finland using summary models. *For. Ecol. Manage.* 259: 524-533. Huang,
498 S.M., S.J. Titus and D.P. Wiens 1992. Comparison of Nonlinear Height Diameter Functions
499 for Major Alberta Tree Species. *Can. J. For. Res.* 22:1297-1304.
- 500 Kantola, A., Mäkinen, H. and Mäkelä, A. 2007. Stem form and branchiness of Norway
501 spruce as a sawn timber – predicted by a process-based model. *For. Ecol. Man.* 241:209-222.
- 502 Kimmins, J.P., Blanco, J.A., Seely, B. and Welham, C. 2008. Complexity in modeling forest
503 ecosystems; how much is enough? *For. Ecol. Man.* 256: 1646-1658.
- 504 Korzhukin, M. D., Ter-Mikaelian, M.T. and Wagner, R. G. 1996. Process versus empirical
505 models: which approach for ecosystem management. *Can. J. For. Res.* 26: 879-887.
- 506
- 507 Landsberg, J.J. and Waring, R.H. 1997. A generalised model of forest productivity using
508 simplified concepts of radiation-use efficiency, carbon balance and partitioning. *For. Ecol.*
509 *Man.* 95: 209 – 228.
- 510 Lo, Y.-H., Blanco, J. A., Kimmins, J.P., Seely, B. and Welham, C. 2011. Linking Climate
511 Change and Forest Ecophysiology to Project Future Trends in Tree Growth: A Review of
512 Forest Models Climate Change. *Research and Technology for Adaptation and Mitigation.*
513 *Edited by J. Blanco and H. Kheradmand, InTech, ISBN 978-953-307-621-8, pp. 63-86.*
- 514 Lonsdale, J., Xenakis, G., Mencuccini, M. and Perks, M. 2015. A comparison of models for
515 quantifying growth and standing carbon in UK Scots pine forests. *iFor.* 8:596-605.
- 516 Mäkelä, A., Pulkkinen, M. and Mäkinen, H. 2016. Bridging empirical and carbon-balance
517 based forest site productivity – Significance of below-ground allocation. *For. Ecol. Man.* 372:
518 64-77.
- 519 Mäkelä, A. 1997. A Carbon Balance Model of Growth and Self-Pruning in Trees Based on
520 Structural Relationships. *For. Sci.* 43:7-24.
- 521 Mäkelä, A., Vanninen, P. and Ikonen, V.-P. (1997). An application of process-based
522 modelling to the development of branchiness in Scots pine. *Silva Fenn.* 31:369-380
- 523 Mäkelä, A. 2002. Derivation of stem taper from the pipe theory in a carbon balance
524 framework. *Tree Physiol.* 22:891-905.
- 525 Mäkelä, A., Landsberg, J., Ek, A.R., Burk, T.E., Ter-Mikaelian, M., Ågren, G.I., Chadwick,
526 D.O. and Puttonen, P. 2000. Process-based models for forest ecosystem management:
527 current state of the art and challenges for practical implementation. *Tree Physiol.* 20:289-298.
- 528 Mäkelä, A. and Mäkinen, H. 2003. Generating 3D sawlogs with a process-based growth
529 model. *For. Ecol. Man.* 184:337-354.

- 530 Mäkelä, A. and Sievänen, R. 1992. Height growth strategies in open-grown trees. *J. Theor.*
531 *Biol.* 159:443-467.
- 532 Mäkinen, H. and Isomäki, A. 2004. Thinning intensity and growth of Norway spruce stands
533 in Finland. *Forestry* 77: 349 – 364.
- 534 Mäkinen, H. and Isomäki, A. 2004. Thinning intensity and long-term changes in increment
535 and stem form of Norway spruce trees. *For. Ecol. Man.* 201: 295 – 309.
- 536 Mäkinen, H., Isomäki, A. and Hongisto, T. 2006. Effect of half-systematic and systematic
537 thinning on the increment of Scots pine and Norway spruce in Finland. *Forestry* 79: 103 –
538 121.
- 539 Mäkipää, R., Linkosalo, T., Komarov, A., Mäkelä, A. 2015. Mitigation of climate change
540 with biomass harvesting in Norway spruce stands — are harvesting practices carbon neutral?
541 *Can. J. For. Res.* 45: 217–225.
- 542 Marklund, L.G. 1988. Biomass functions for pine, spruce and birch in Sweden. *Swed. Univ.*
543 *of Agric. Sciences, Dep. of For. Surv., Report, 45, p. 73* (In Swedish with English summary).
- 544 Minunno, F., Xenakis, G., Perks, M.P. and Mencuccini, M. 2010. Calibration and validation
545 of a simplified process-based model for the prediction of the carbon balance of Scottish Sitka
546 spruce (*Picea sitchensis*) plantations. *Can. J. For. Res.* 40: 2411 – 2426.
- 547 Minunno, F., Peltoniemi, M., Launiainen, S., Aurela, M., Kolari, P., Lindroth, A., Lohela,
548 A., Minkkinen, K., Mäkelä, A., 2016. A simplified gross primary production and
549 evapotranspiration model for Boreal coniferous forests – is a generic calibration
550 sufficient? Manuscript submitted to *Ecological Modelling*.
- 551
- 552 Medlyn B.E., Berbigier P., Clement R., Grelle A., Loustau D., Linder S., Wingate L., Jarvis
553 P.G., Sigurdsson B.D., McMurtrie R.E. 2005. Carbon balance of coniferous forests growing
554 in contrasting climates: Model-based analysis. *Agr. For. Met.* 131: 97-124.
- 555 Monserud, R.A. 2003. Evaluating forest models in a sustainable forest management context.
556 *For. Biom., Model. Info.* 1: 35 – 47.
- 557 Niinimäki, S., Tahvonen, O., Mäkelä, A. and Linkosalo, T. 2013. On the economics of
558 Norway spruce stands and carbon storage. *Can. J. For. Res.* 43: 637 – 648.
- 559 Niinimäki, S., Tahvonen, O. and Mäkelä, A. 2012. Applying a process-based model in
560 Norway spruce management. *For. Ecol. Man.* 265: 102 – 115.
- 561 Pérez-Cruzado, C., Muñoz-Sáez, F., Basurco, F., Riesco, G. and Rodríguez-Soalleiro, R.
562 2011. Combining empirical models and the process-based model 3-PG to predict *Eucalyptus*
563 *nitens* plantations growth in Spain. *For. Ecol. Man.* 262: 1067 – 1077.
- 564 Petersson, H. 1997. Functions for predicting crown height of *Pinus sylvestris* and *Picea abies*
565 in Sweden. *Scandinavian J. For. Res.* 12: 179 – 188.
- 566 Pietritsch, R., Hasenauer, H. and Pietsch, S.A. 2007. Incorporating forest growth response to
567 thinning within Biome-BGC. *For. Ecol. Man.* 242: 324 – 336.
- 568 Pinjuv, G., Mason, E.G. and Watt, M. 2006. Quantitative validation and comparison of a
569 range of forest growth model types. *For. Ecol. Man.* 236: 37 – 46.

- 570 Pretzsch, H. 2002. Application and evaluation of the growth simulator SILVA 2.2 for forests
571 stands, forest estates and large regions. *Forst. Cent.* 121: 28 – 51.
- 572 Raulier, F., Pothier, D. and Bernier, P.Y. 2003. Predicting the effect of thinning on growth of
573 dense balsam fir stands using a process-based tree growth model. *Can. J. For. Res.* 33: 509 –
574 520.
- 575 Reineke, L.H. 1933. Perfecting a stand-density index for even-aged forests. *J. Agr. Res.* 46:
576 627 – 638.
- 577 Repola, J. 2009. Biomass Equations for Scots Pine and Norway Spruce in Finland. *Silva*
578 *Fenn.* 43: 625 – 647.
- 579 Robinson, A.P. and Ek, A.R. 2003. Description and validation of a hybrid model of forest
580 growth and stand dynamics for the Great Lakes region. *Ecol. Model.* 170: 73 – 104.
- 581 SAS Institute Inc. 2015. SAS® 9.4 Statements: Reference, Fourth Edition. Cary, NC: SAS
582 Institute Inc.
- 583 Seely, B., Welham, C. and Scoullar, K. 2015. Application of a hybrid forest growth model to
584 evaluate climate change impacts on productivity, nutrient cycling and mortality in a montane
585 forest ecosystem. *PlosONE* doi: 10.1371/journal.pone.0135034
- 586 Shinozaki, K., Yoda, K., Hozumi, K. and Kira, T. 1964. A quantitative analysis of plant
587 form-the pipe model theory. I Basic analyses. *Jap. J. Ecol.* 14: 97 – 105.
- 588 Thornley, J.H.M. and Cannell, M.G.R. 2000. Managing forests for yield and carbon storage:
589 a theoretical study. *Tree Physiol.* 20: 477 – 484.
- 590 Valentine, H.T., Mäkelä, A. 2005. Bridging process-based and empirical approaches to
591 modeling tree growth. *Tree Physiol* 25: 769 – 779.
- 592 Valentine H.T., Amateis R.L., Gove H., Mäkelä A. 2013. Crown rise and crown-length
593 dynamics: application to loblolly pine. *Forestry.* 86:371-375
- 594 Valentine H.T., Mäkelä A., Green E.J., Amateis R.L., Mäkinen H., Ducey M.J. 2012. Models
595 relating stem growth to crown length dynamics: application to loblolly pine and Norway
596 spruce. *Trees Structure and Function.* 26:469–478
- 597 Van Oijen, M., Reyer, C., Bohn, F.J., Cameron, D.R., Deckmyn, G., Felchsig, M., Härkönen,
598 S., Hartig, F., Huth, A., Kiviste, A., Lasch, P., Mäkelä, A., Mette, T., Minunno, F., Rammer,
599 W. 2013. Bayesian calibration, comparison and averaging of six forest models, using data
600 from Scots pine stands across Europe. *For. Ecol. Man.* 289: 255–268.
- 601 Vanclay, J.K. and Skoovsgard, J.P. 1997. Evaluating forest growth models. *Ecol. Model.* 98:
602 1 – 12.
- 603 Vuokila, Y. and Väliäho, H. 1980. Viljeltyjen havumetsiköiden kasvatusmallit. Summary:
604 Growth and yield models for conifer cultures in Finland. *Comm. Inst. For. Fenn.* 99(2). 271 s.
- 605 Zhou, X., Dang, Q.-L., Chen, J. 2005. Predicting forest growth and yield in northeastern
606 Ontario using the process-based model of TRIPLEX1.0. *Can. J. For. Res.* 35: 2268 – 2280.
- 607

608 **Figure captions**609 **Fig. 1.** The structure of PipeQual model.

610 **Fig. 2.** An example of the estimation of the crown base height for the initial stage of
611 PipeQual model by empirical model of Petersson (1997). Dots are the observed crown base
612 heights and squares the predictions of model. Line is the result of smoothing (eq. 8). In this
613 specific stand and plot (located in Keuruu, ID = Vh002, plot = 2), parameter values of
614 Gompertz function were $A = 4.2129$, $\beta = 1.0740$ and $\kappa = 0.1559$.

615 **Fig. 3.** The tree and stand level relationships between simulated and observed variables. a)
616 stem diameter at breast height (cm), b) dominant height (m), c) stand density (stems ha⁻¹) d)
617 basal area (m² ha⁻¹), e) stand volume (m³ ha⁻¹). Each dot represents one permanent sample
618 plot (PSP) and different colors specify different stands. Continuous line is 1:1 line and dashed
619 line the linear regression curve obtained with coefficients in subfigures.

620 **Fig. 4.** Simulated vs observed diameter distribution. The average of all stand distributions in
621 the last measurement.

622 **Fig. 5.** Residuals (observation - simulation) of a) annual volume increment (volume
623 production), b) annual net increment (rate of change of standing volume), and c) annual drain
624 (harvested + mortality).

625 **Fig. 6.** Dynamics of stand basal area (a-c) and stand volume (d-f) over stand age in three
626 stands (top row Heinola, ID = Vh012, middle Hauho, ID = Ha001, bottom Punkaharju, ID =
627 Pu041). Continuous lines with symbols are the observations, dashed lines with symbols are
628 the PipeQual predictions. Symbols depict the treatment intensities of the experiments.

629 **Fig 7.** Stem form (height to diameter ratio, H/D). a) The relationship between simulated and
630 observed H/D in the whole dataset. Each dot represents one permanent sample plot (PSP) and
631 different colors specify different stands. Continuous line is 1:1 line and the dashed line the
632 linear regression curve obtained with coefficients in the figure. b) The change of H/D over
633 stand age in one stand (located in Hauho, ID = Ha001) with different treatments. Continuous
634 lines with symbols are the observations and the dashed lines with symbols are the PipeQual
635 predictions. Symbols depict the treatment intensities of the experiment.

636 **Fig.8.** Height to crown base. a) The relationship between simulated and observed height to
637 crown base in the whole dataset. Each dot represents one permanent sample plot (PSP) and
638 different colors specify different stands. Continuous line is 1:1 line and the dashed line the
639 linear regression curve obtained with coefficients in the figure. b) Height to crown base over
640 stand age in one stand (located in Heinola, ID = Nyn3). Continuous lines with symbols are
641 the observations and the dashed lines with symbols are the PipeQual predictions. Symbols
642 depict the treatment intensities of the experiment.

643 **Fig. 9.** Stand leaf area index (LAI) over stand age. a) Lines represent the means of different
644 models over all stands in the PSP dataset. Bands are ± 1 standard deviation. b) foliage
645 dynamics in one stand on the unthinned plot (located in Heinola, ID = Vh001) predicted by
646 different models.

647

648 **Table legends**

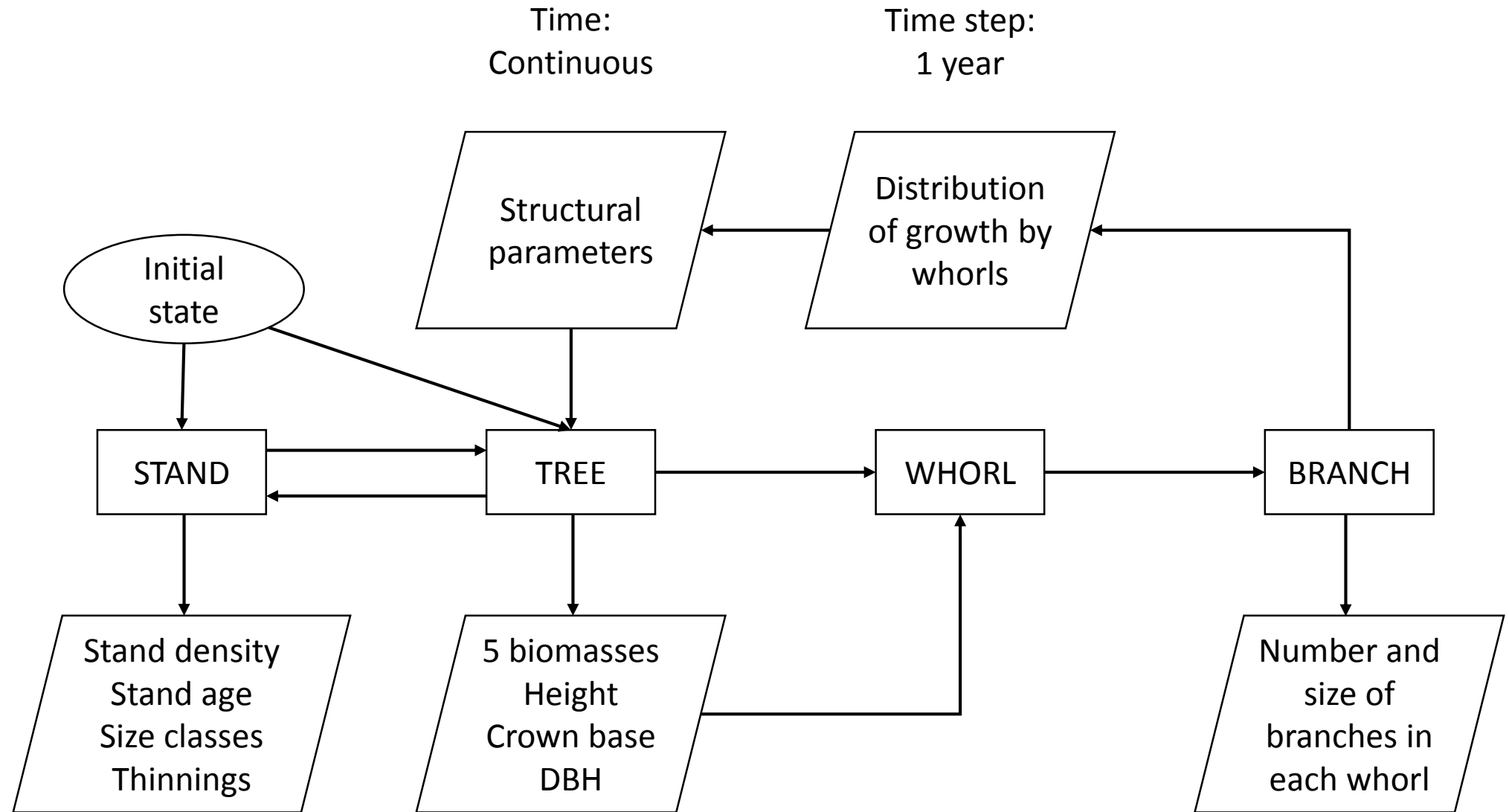
649 **Table 1.** Stand characteristics of the permanent sample plot (PSP) datasets.

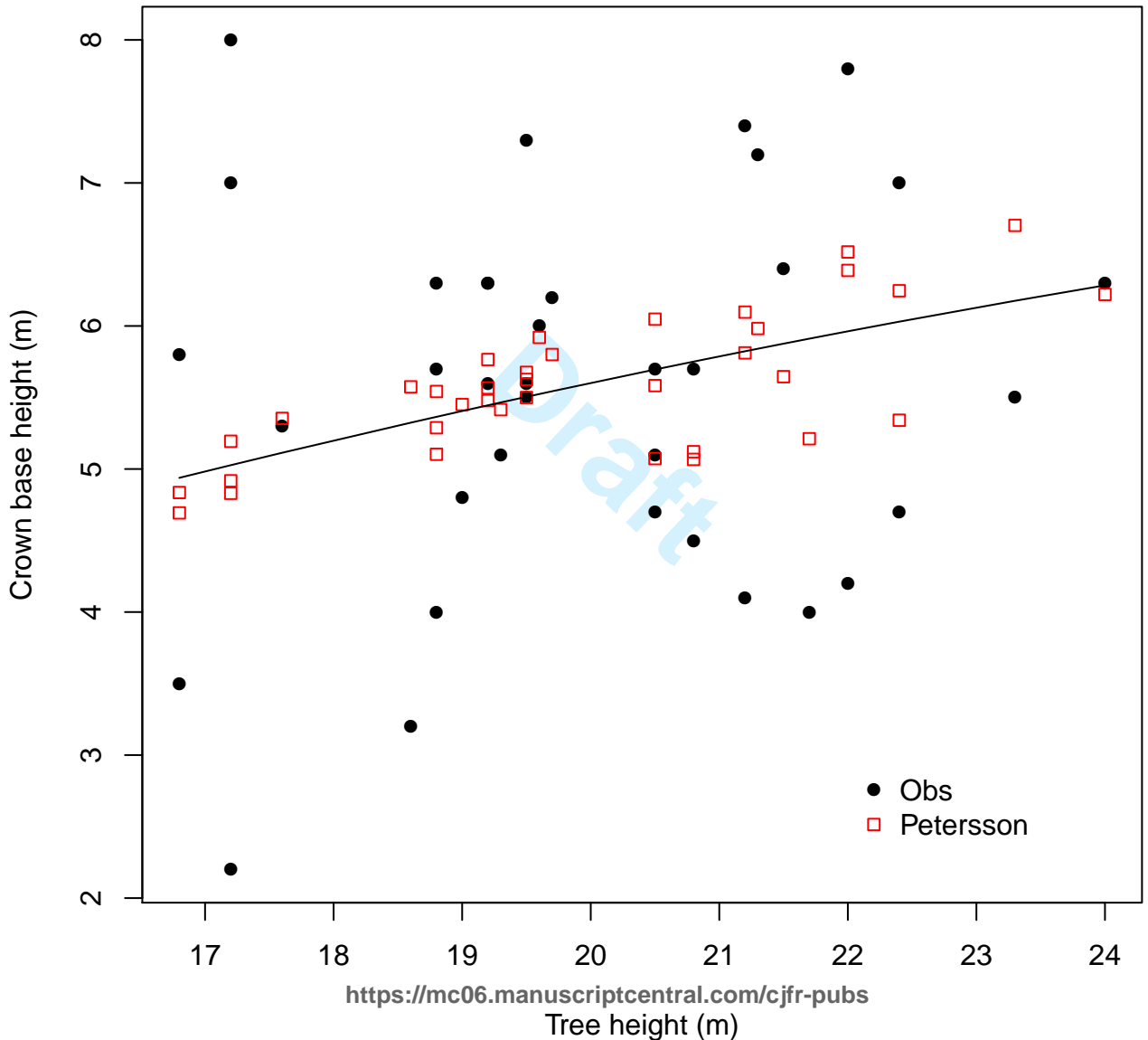
650 **Table 2.** Independent variables and the parameter estimates used in the prediction of crown
651 base height (Eq. 7).

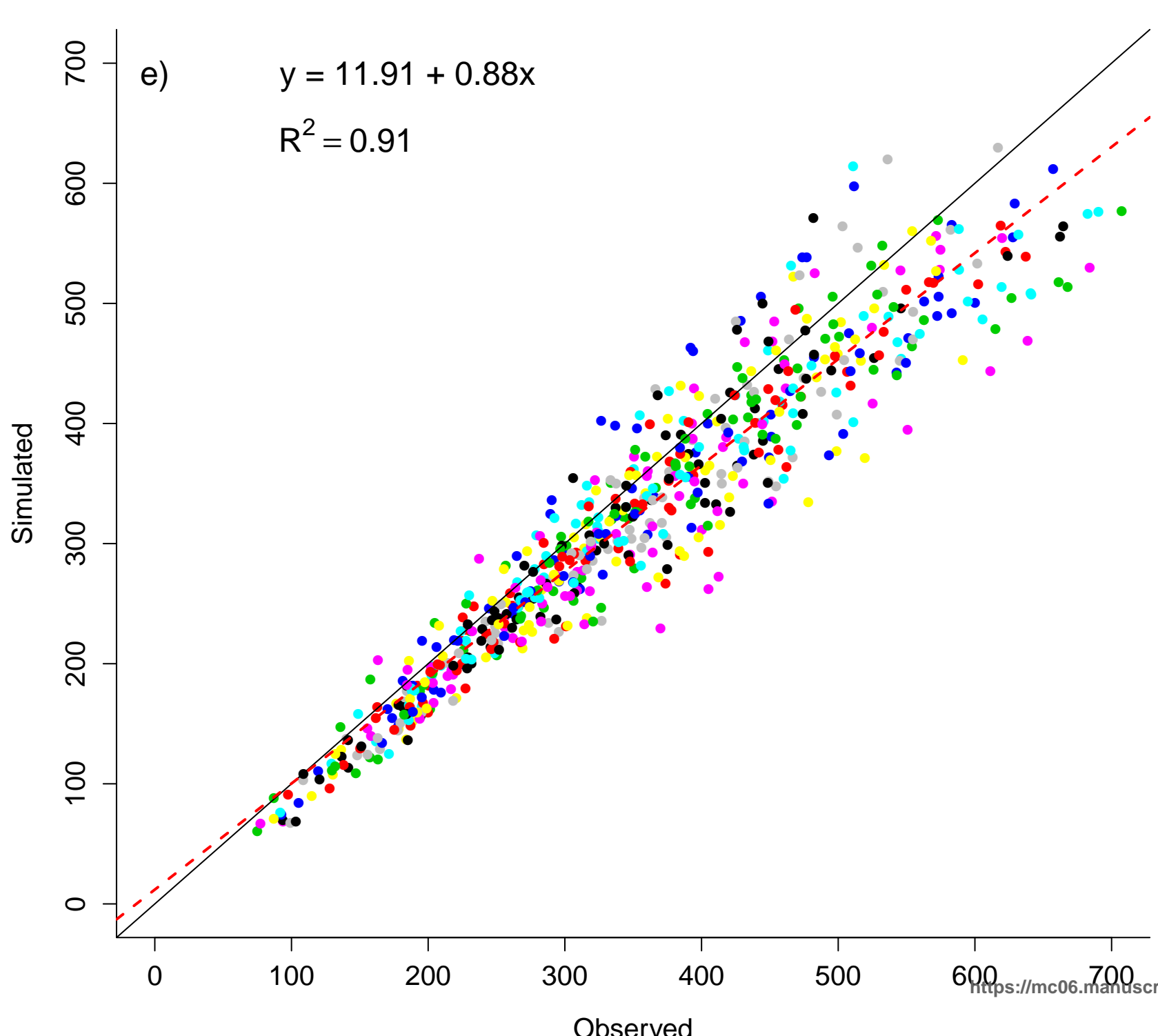
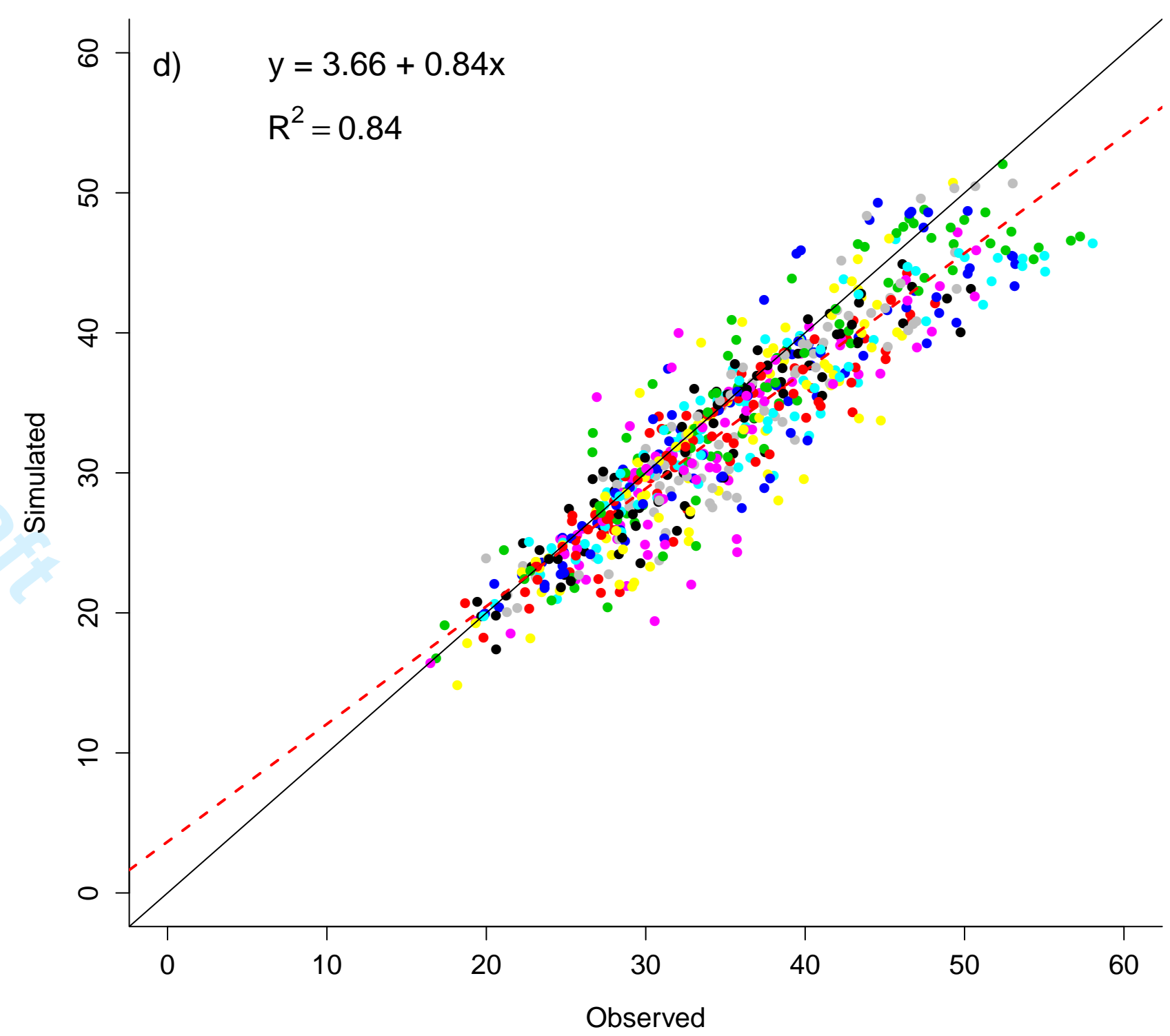
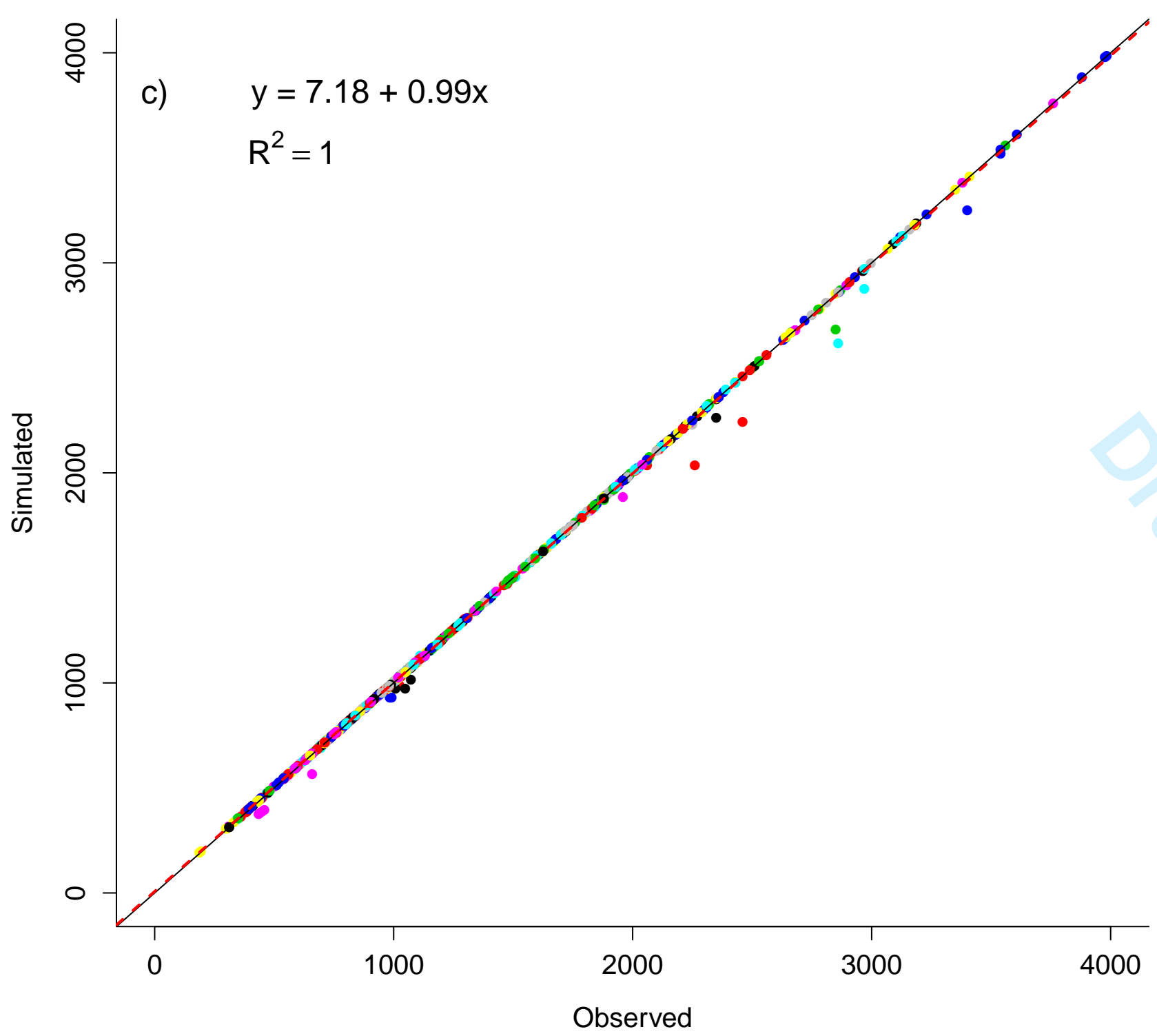
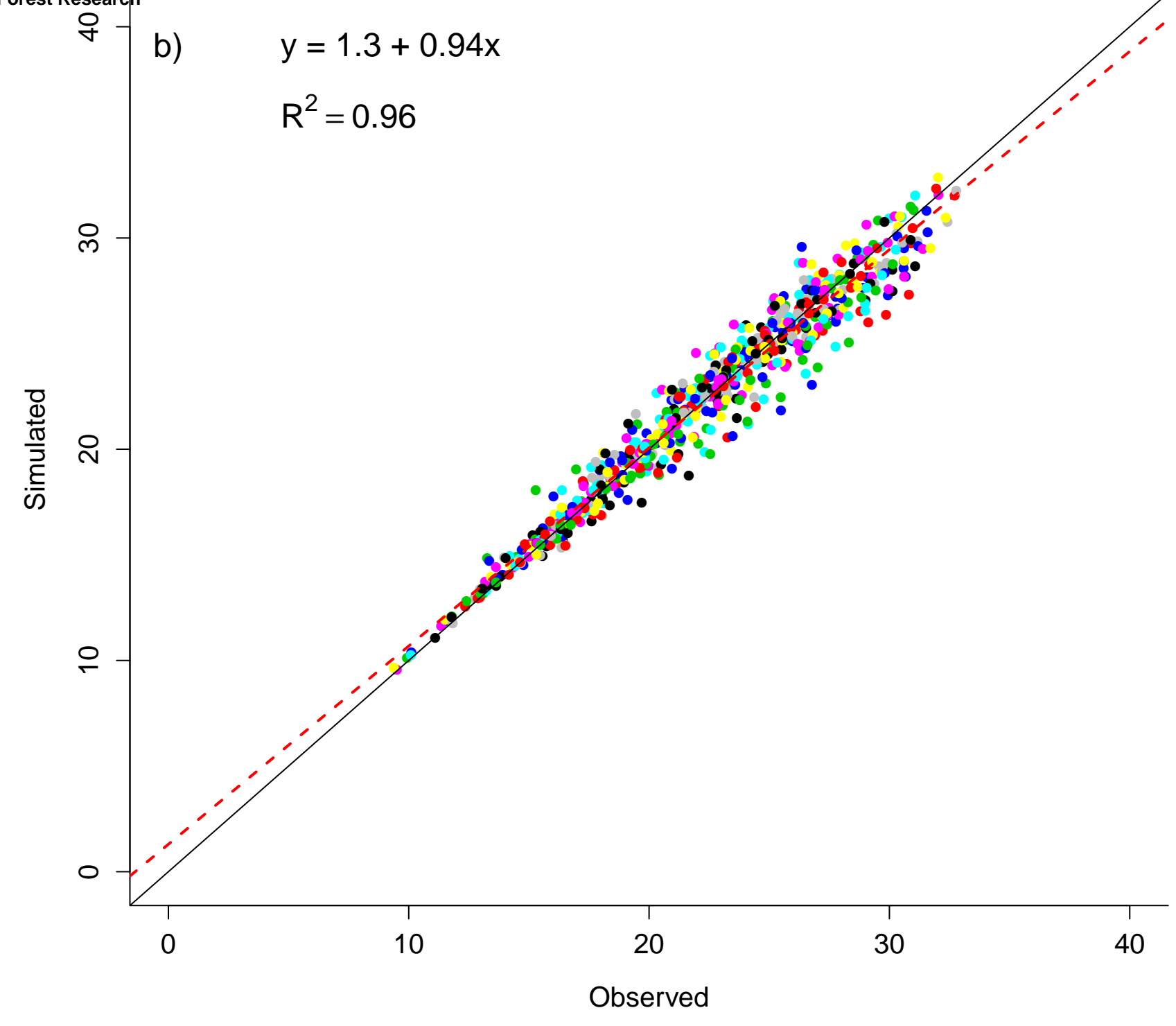
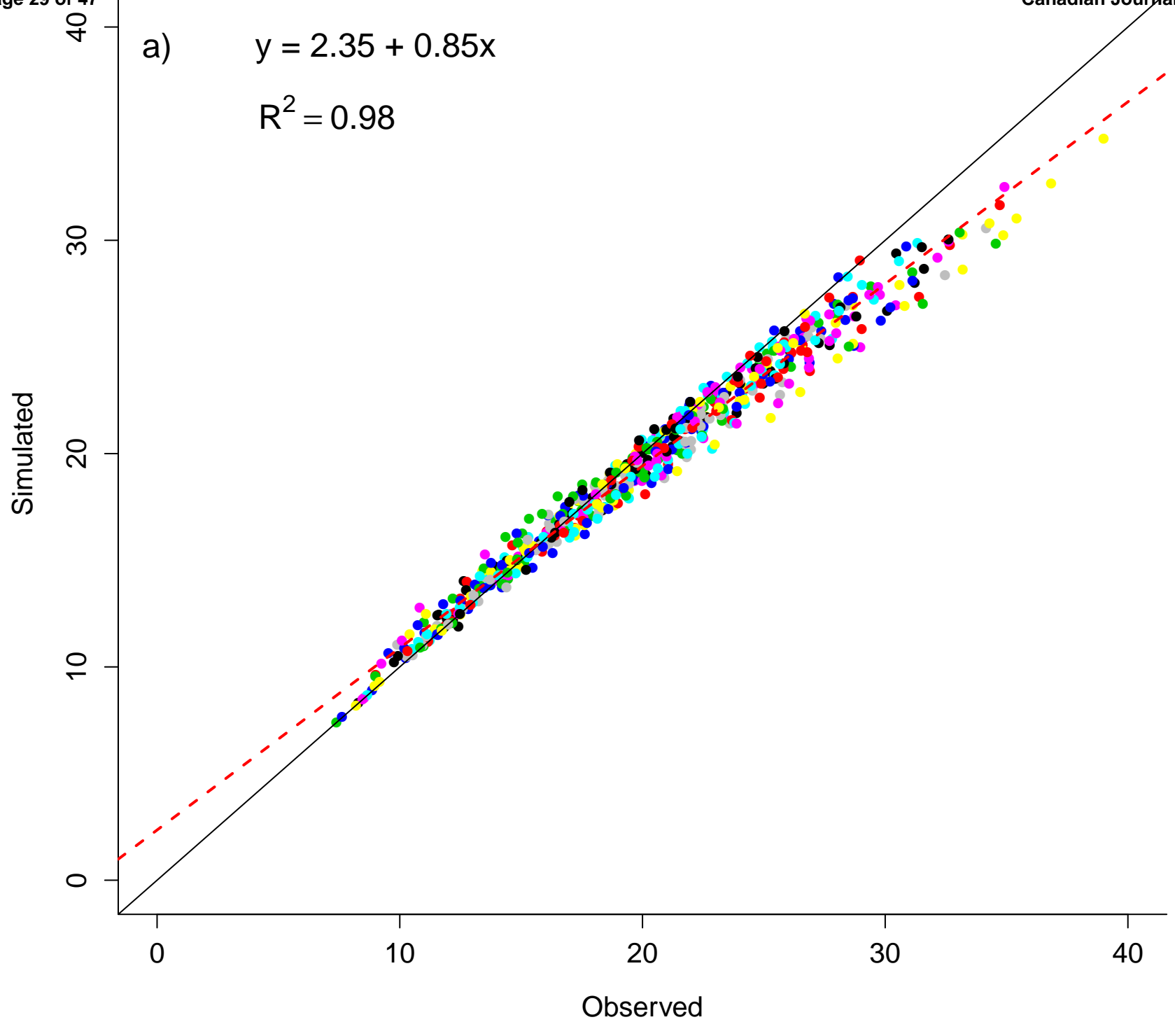
652 **Table 3.** Validation metrics. AMB = absolute model bias (negative values are overestimates,
653 positive underestimates, cm in diameter, m in dominant height, $\text{m}^2 \text{ha}^{-1}$ in basal area, and m^3
654 ha^{-1} in stand volume), RMB = relative model bias, EF = modeling efficiency.

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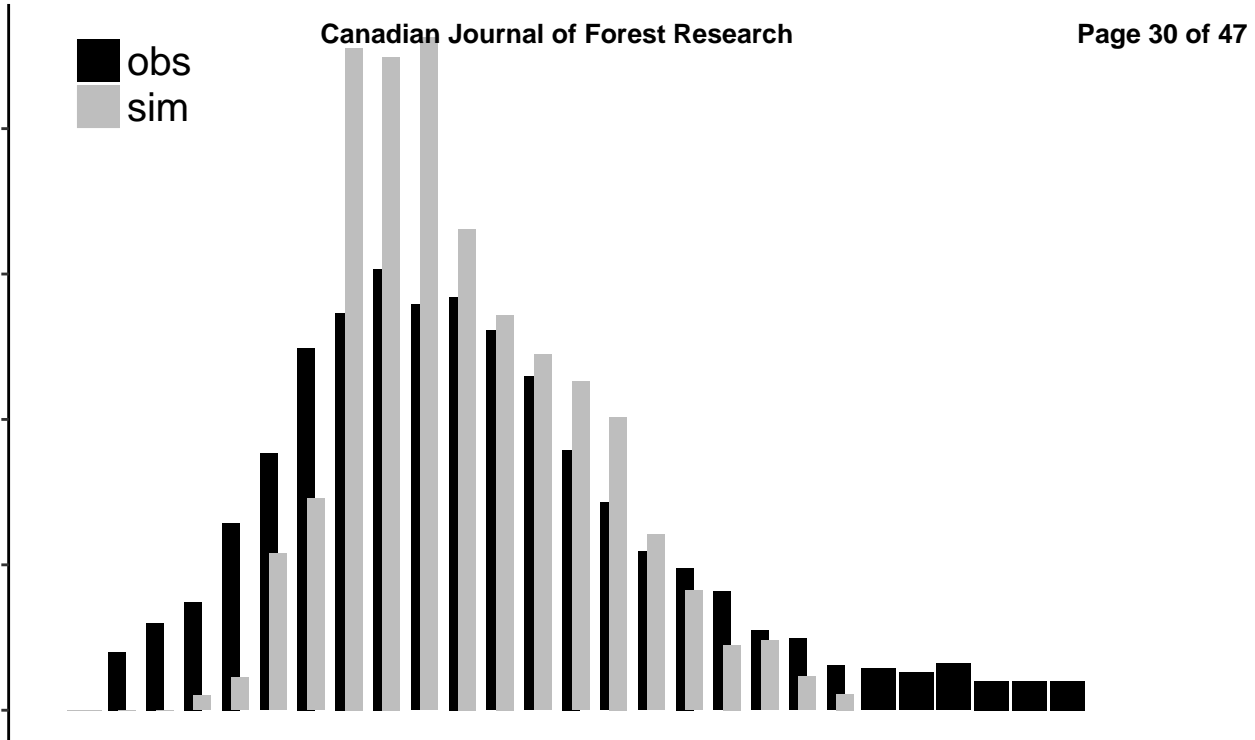
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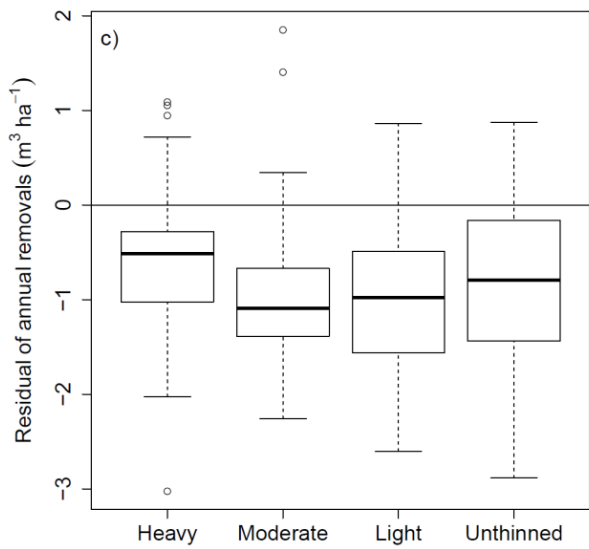
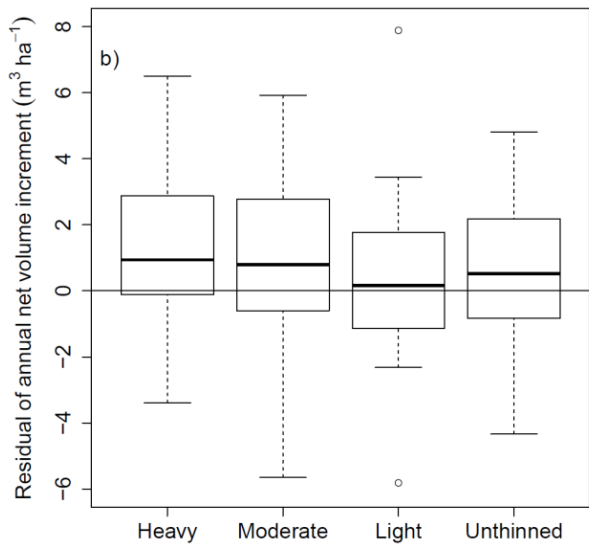
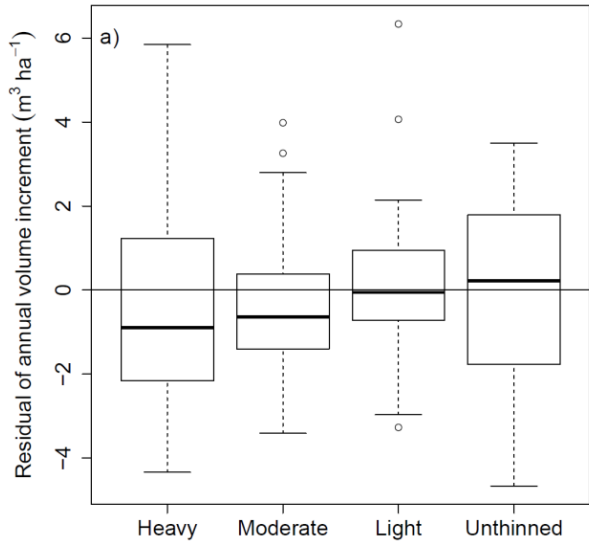
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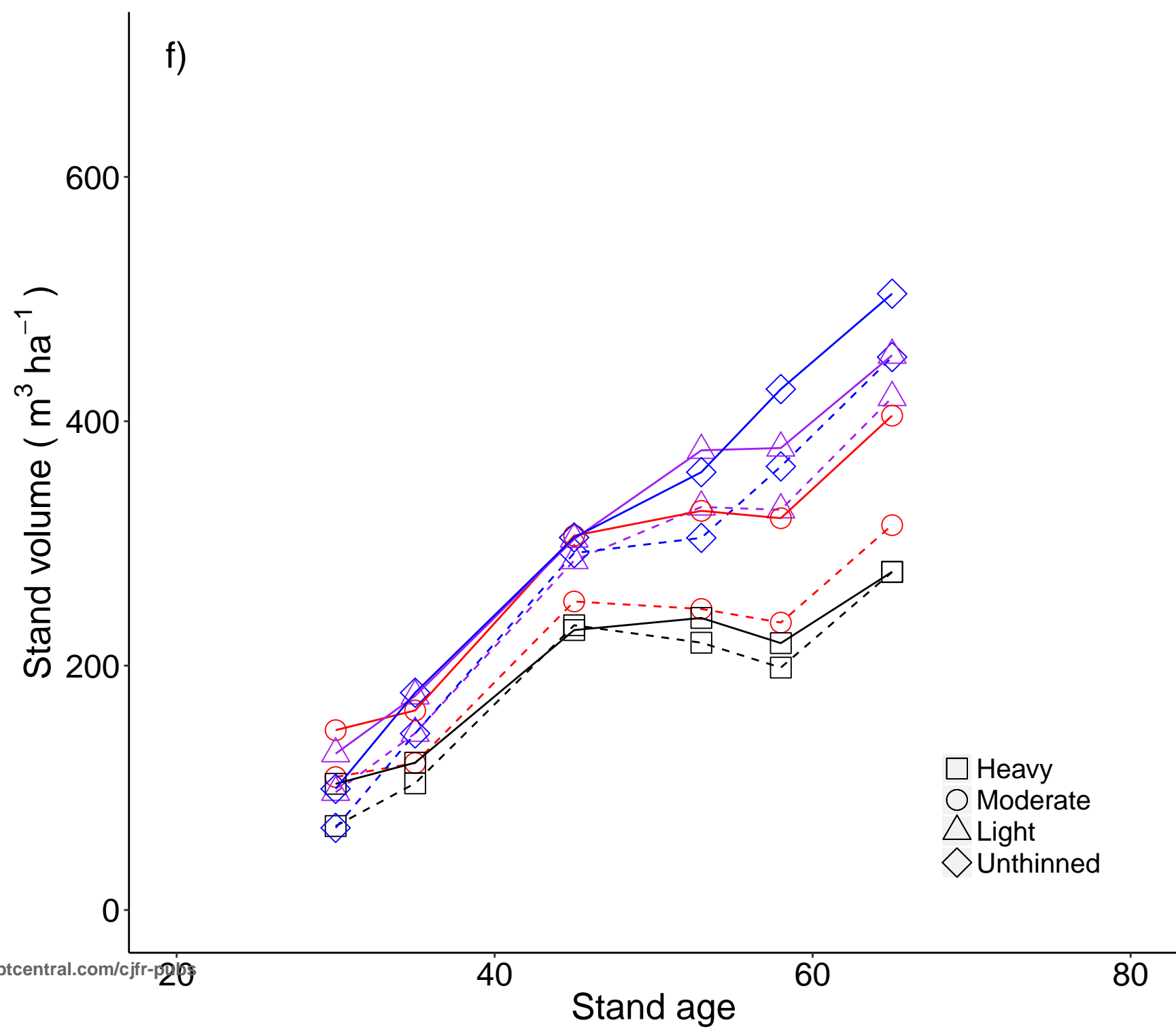
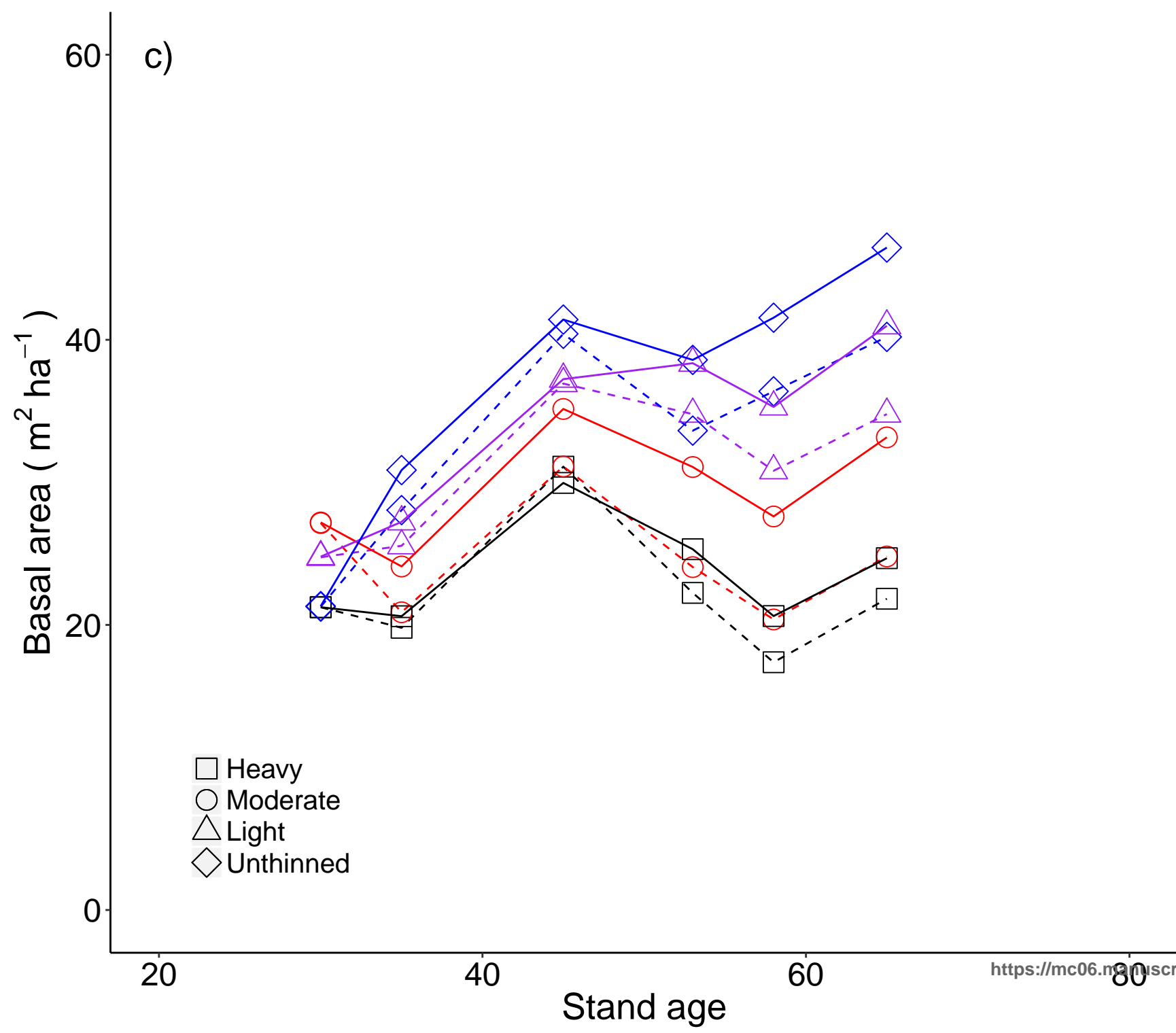
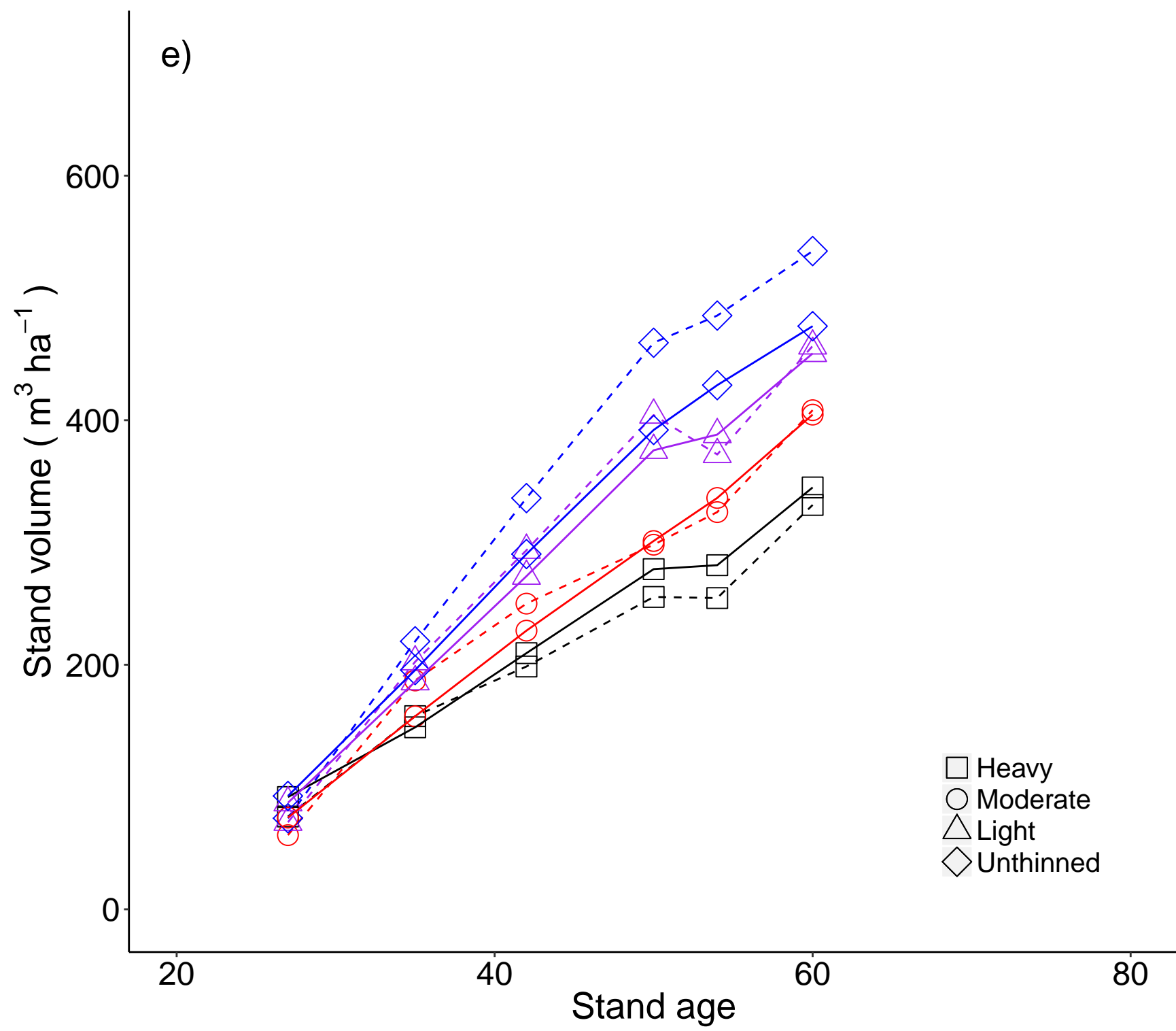
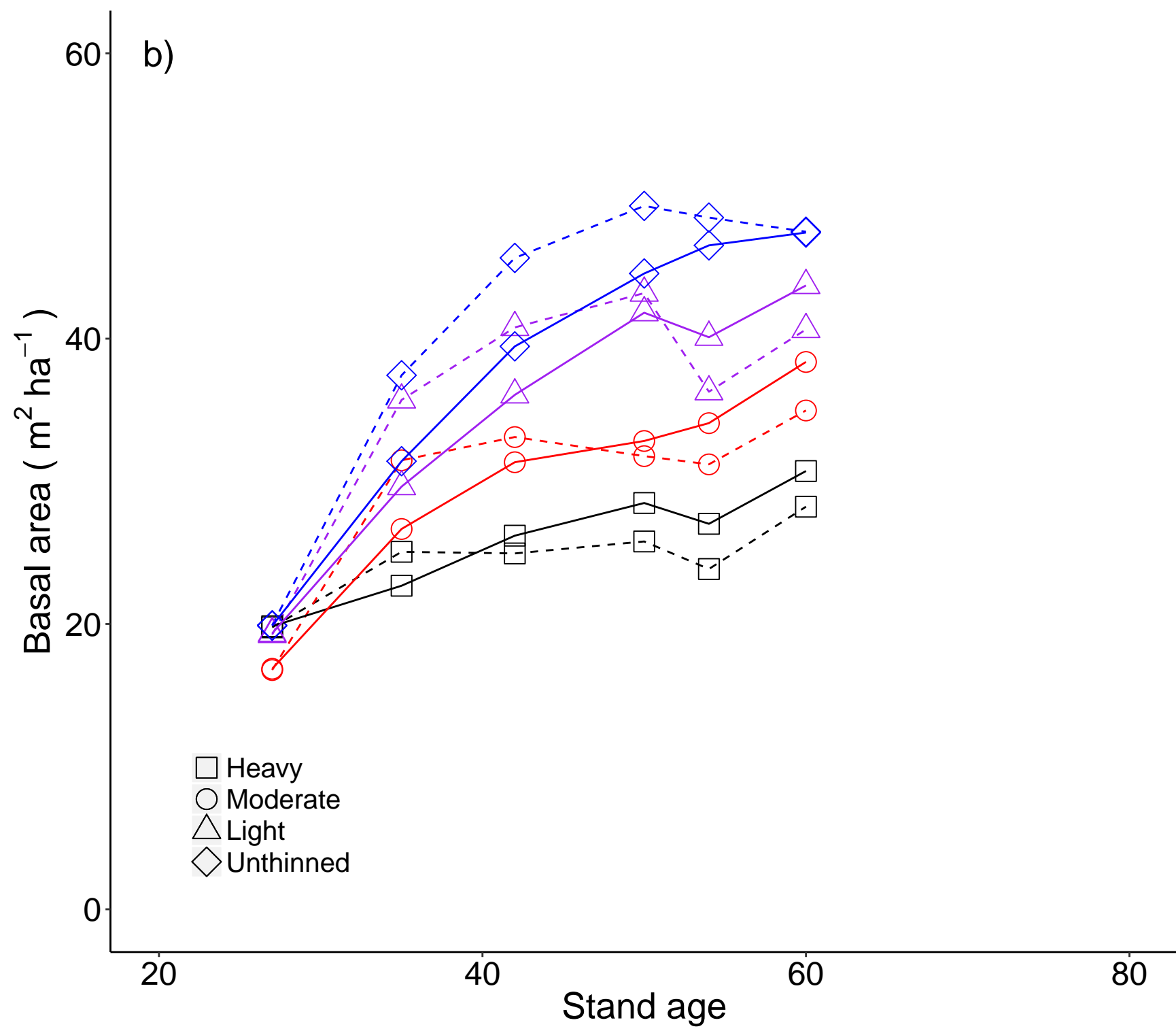
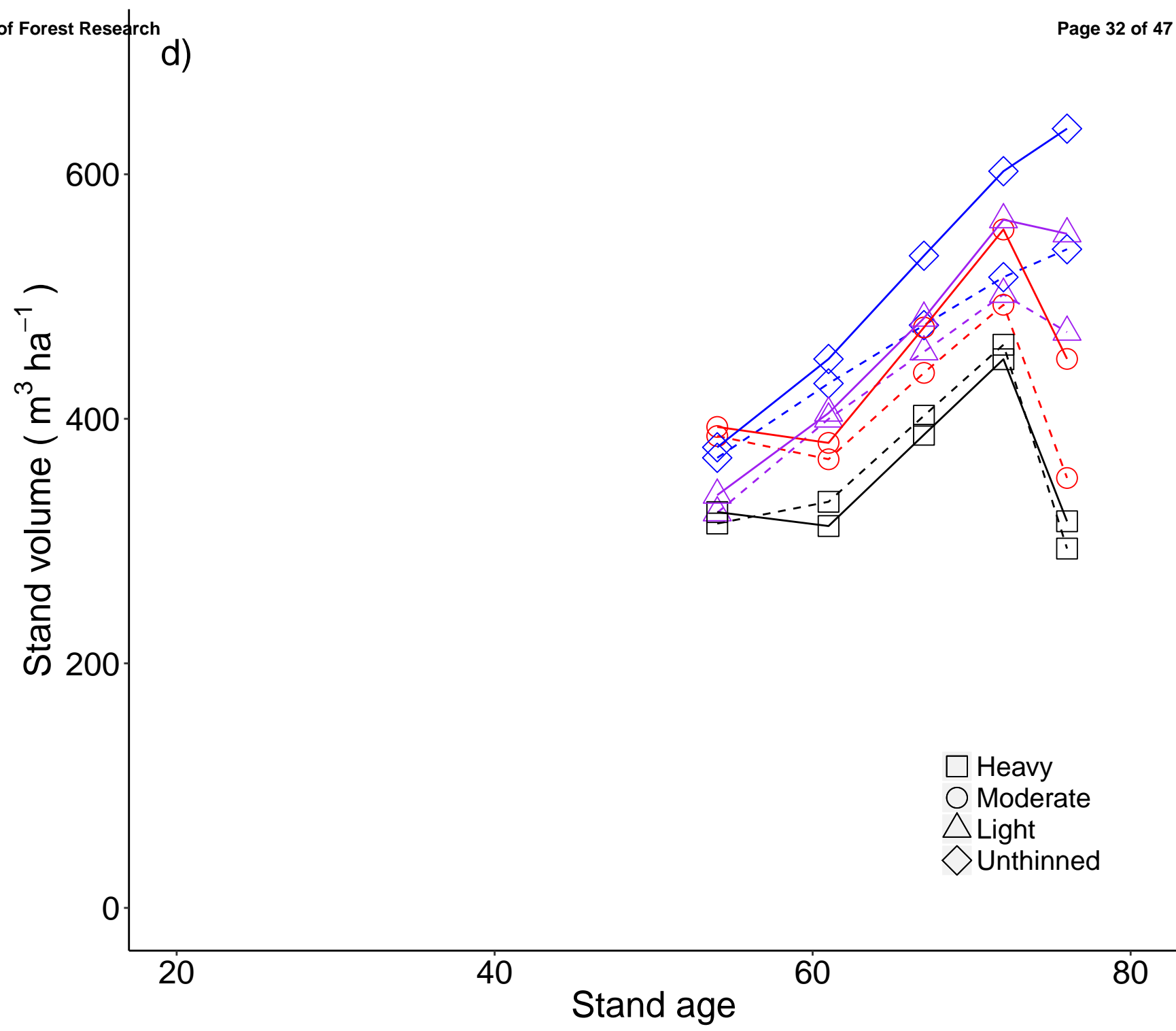
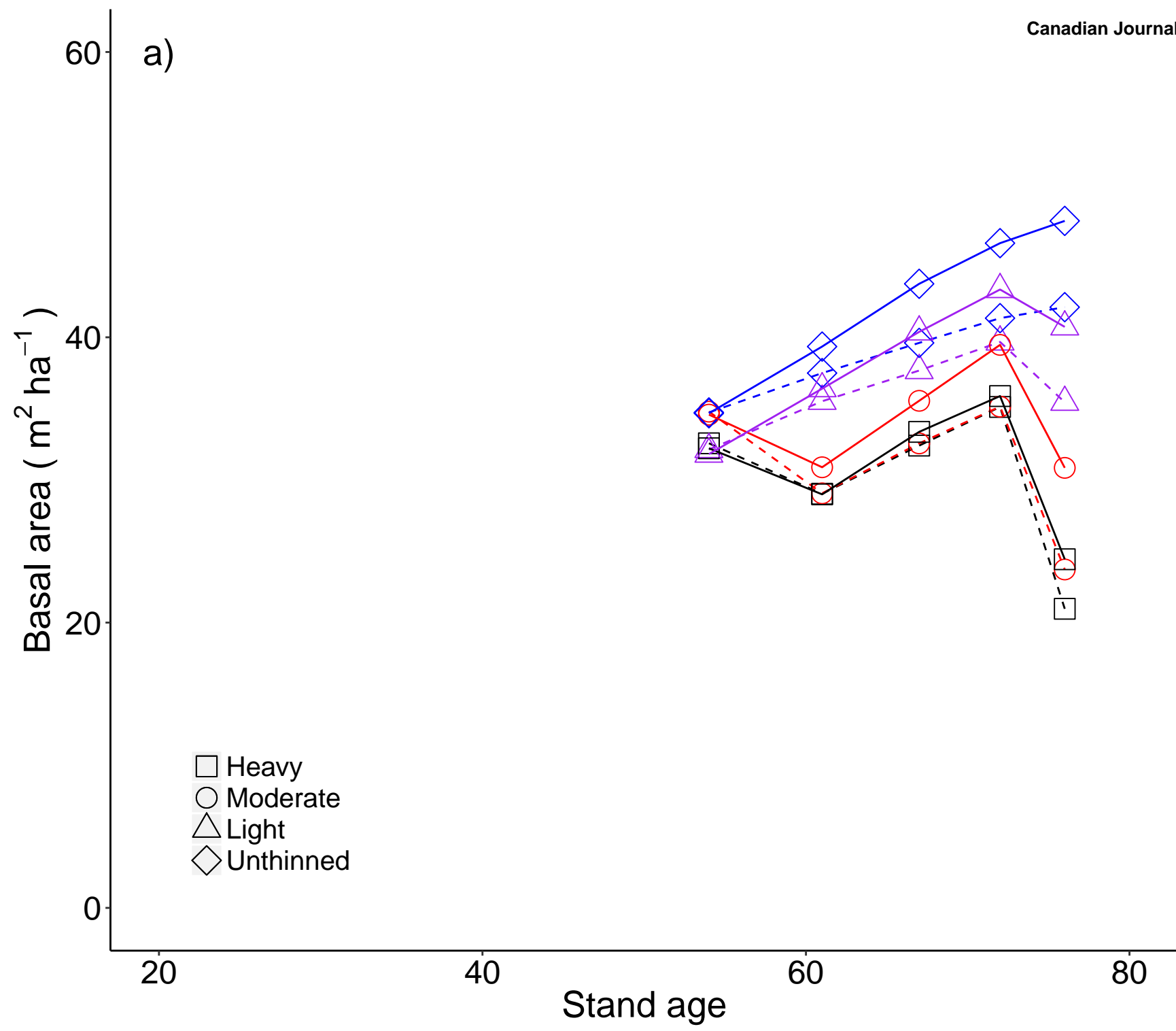
Diameter (cm)

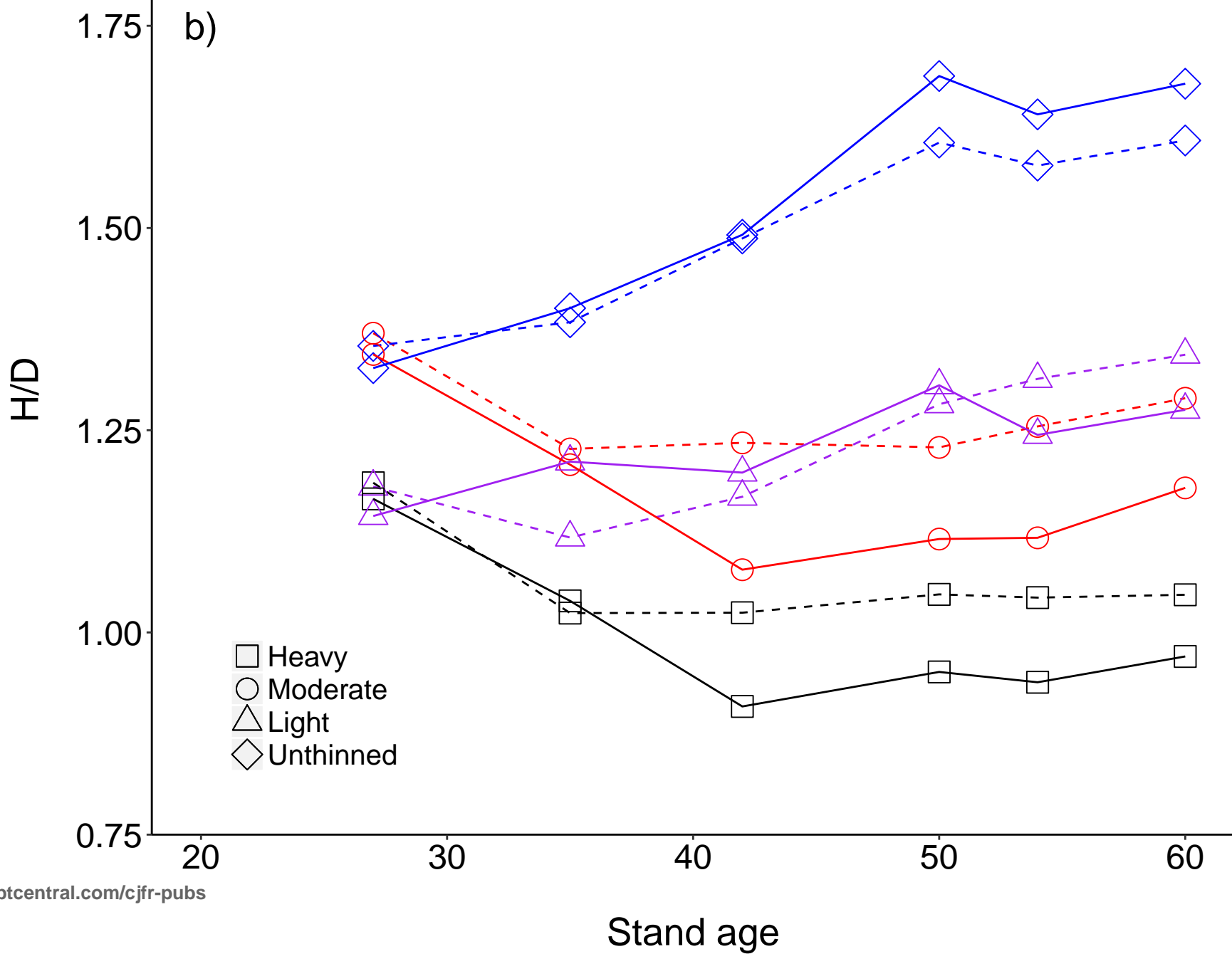
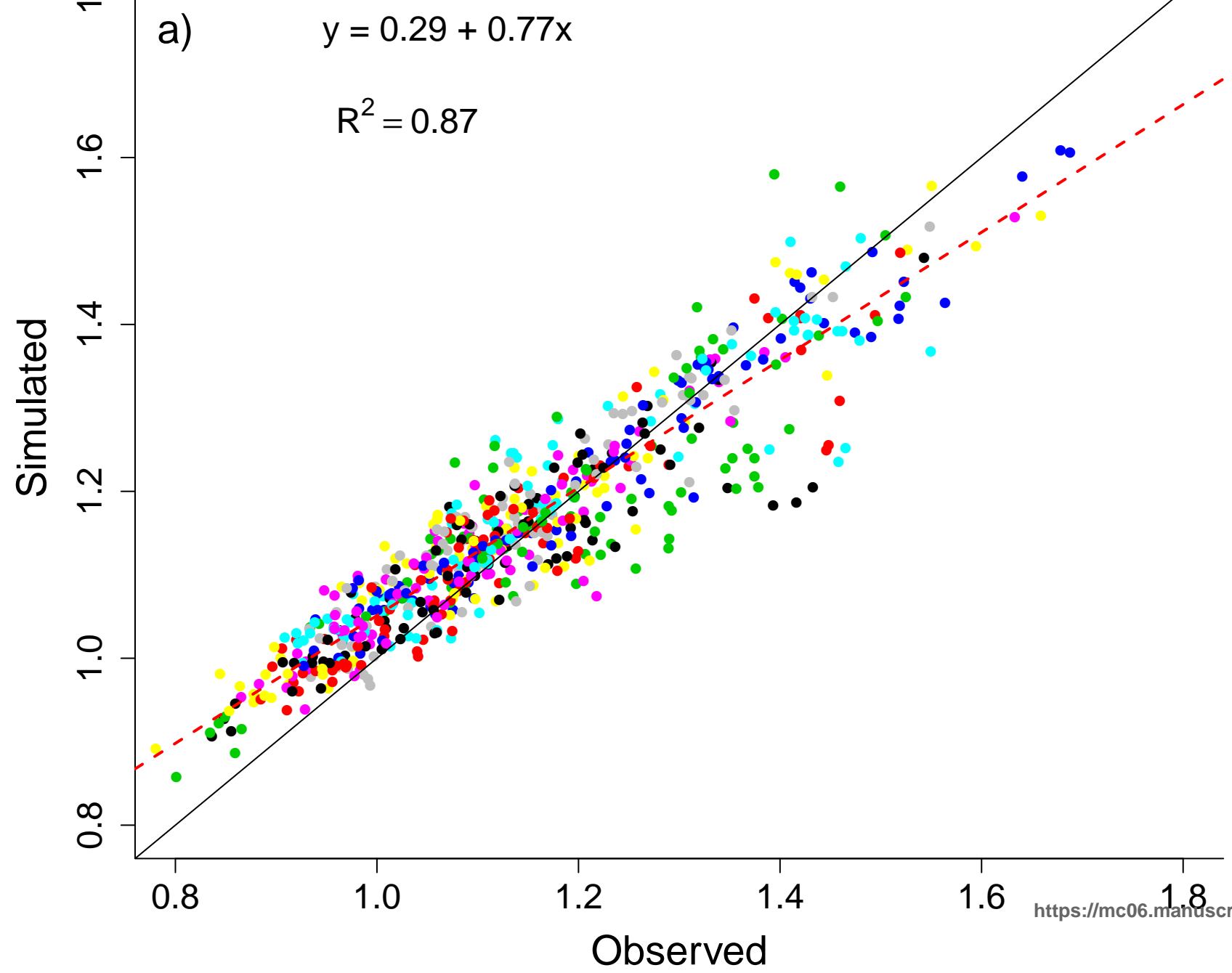
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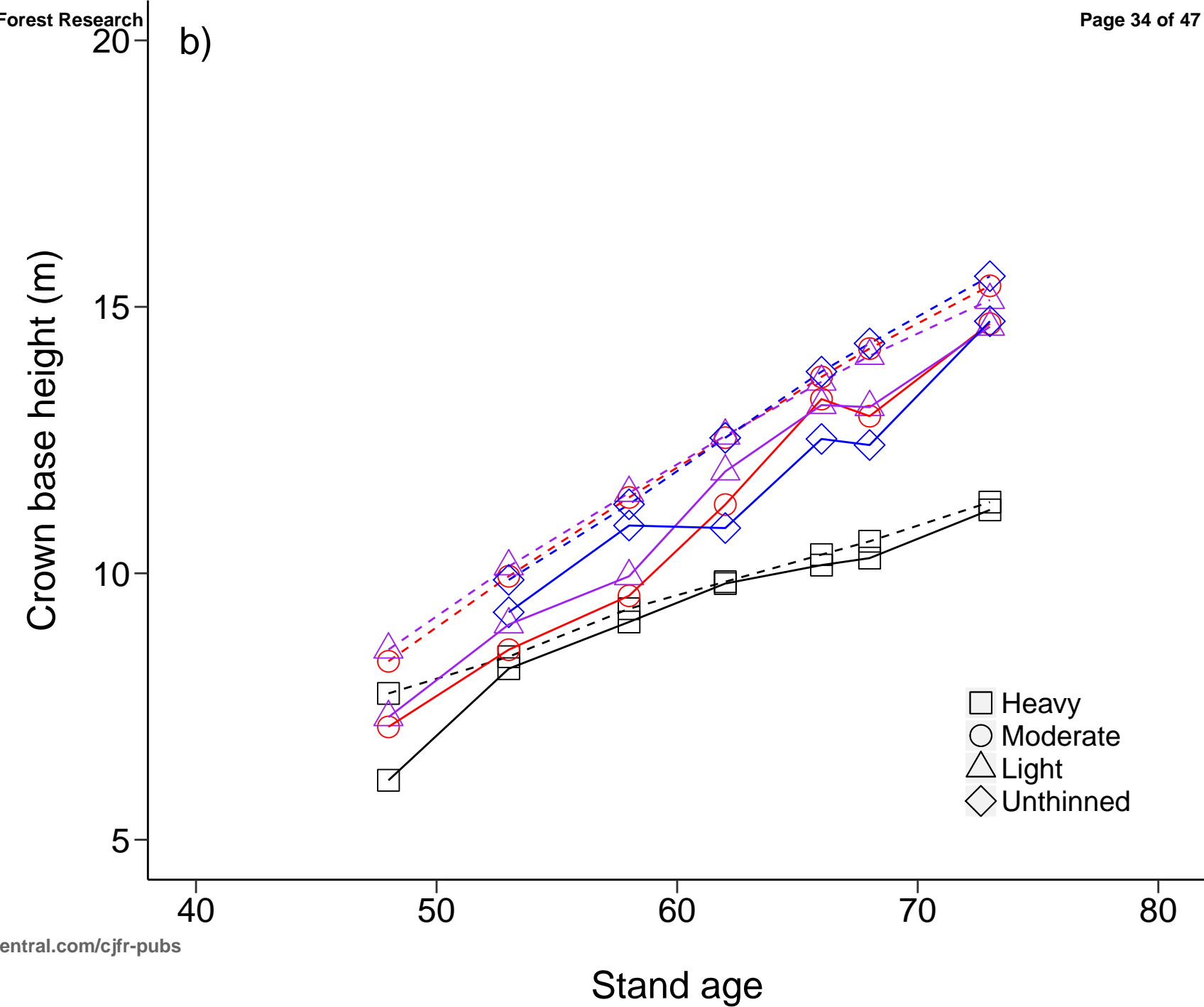
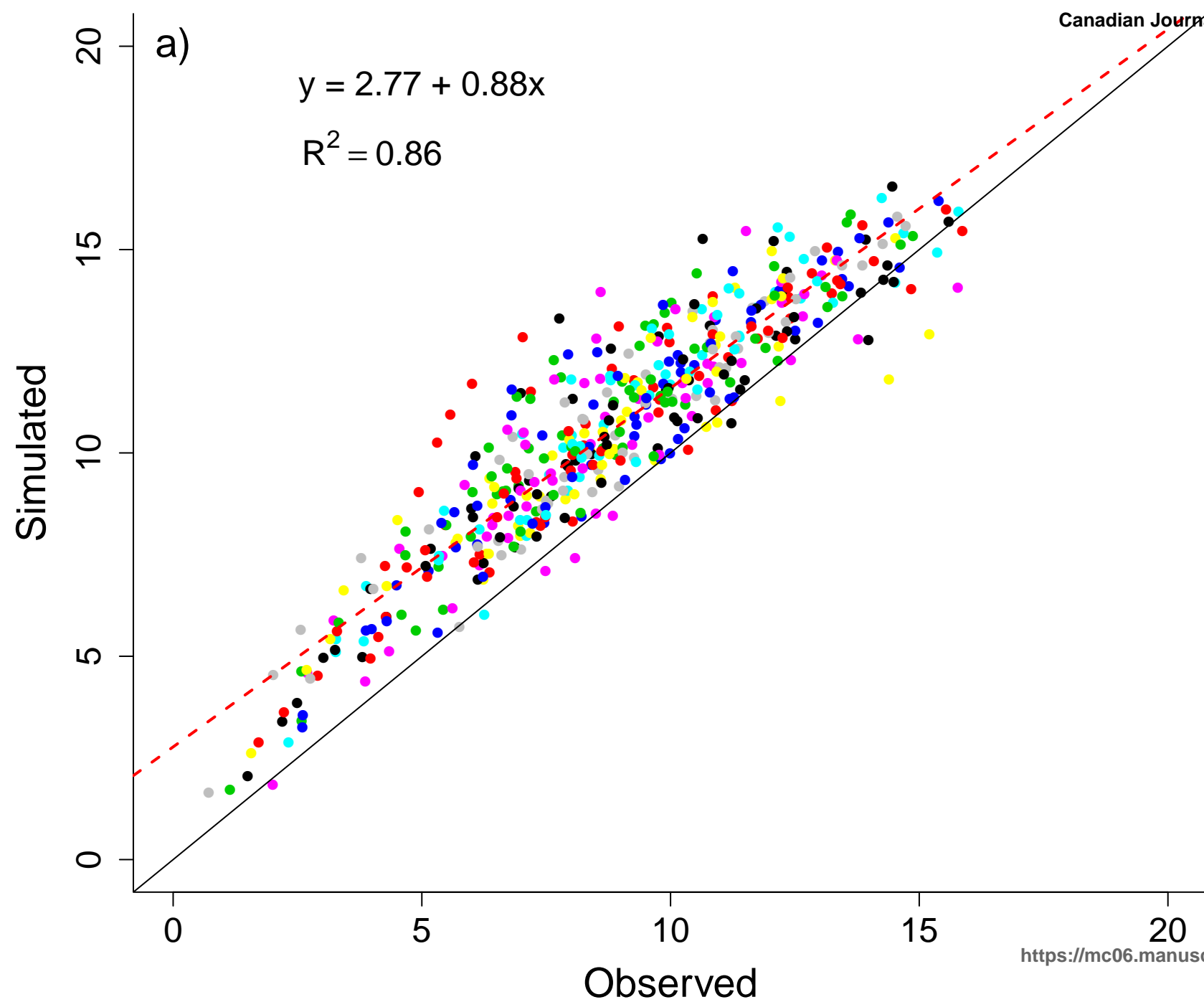


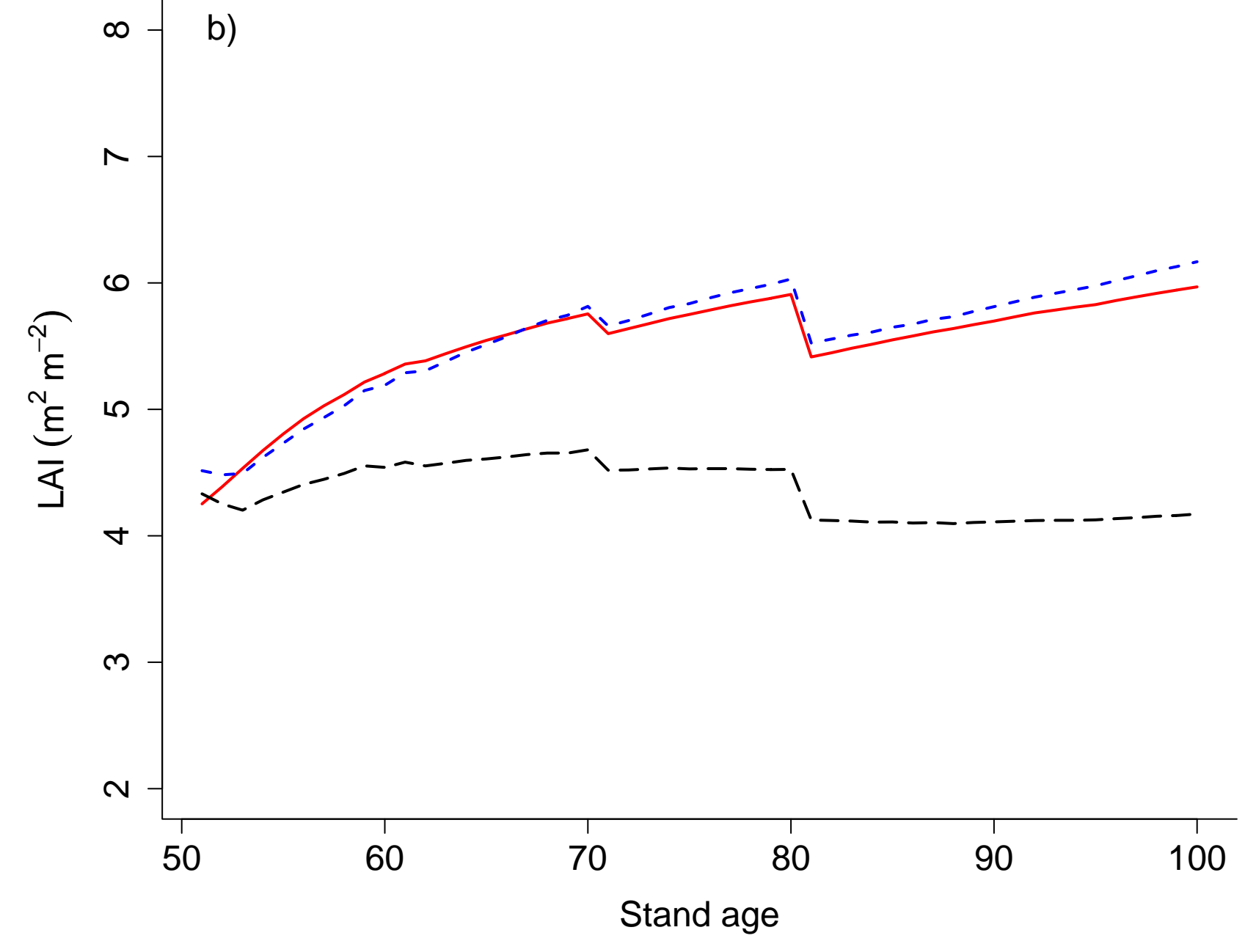
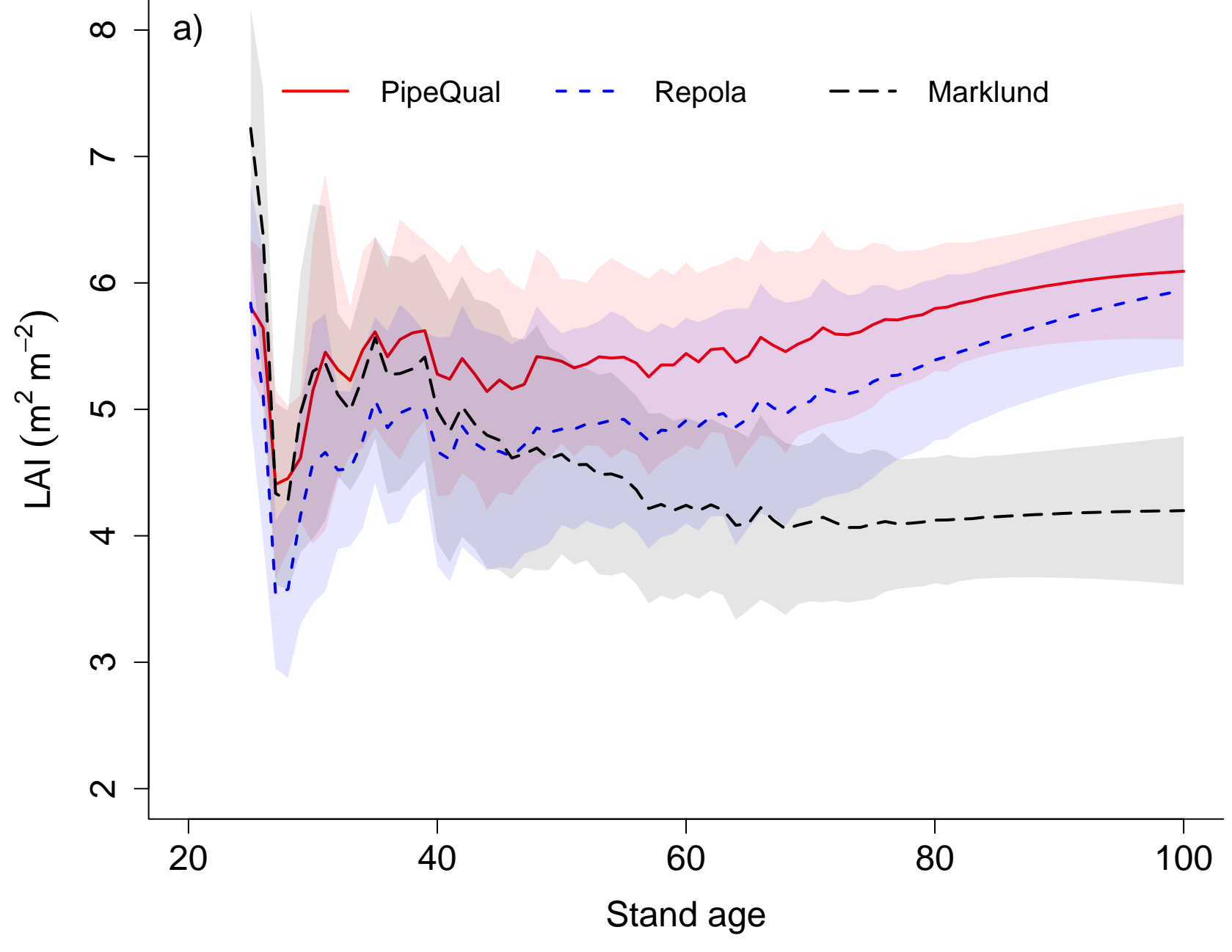


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Table 1. Stand characteristics of the permanent sample plot (PSP) datasets.

Experiment	ID	N of plots	Forest type	H ₁₀₀	Year of planting	Establishment of plots	Last measured	At establishment	
								H _{dom} (m)	N of stems
<i>Harkas</i>									
1	Vh001	6	OMT	34.4	1918	1970	1998	24.3	1072
2	Vh002	3	OMT	31.6	1925	1979	1998	16.9	1404
3	Vh005	10	MT	28.9	1931	1971	1994	13.5	2212
4	Vh009	4	OMT	31.1	1931	1973	1998	17.0	1856
5	Vh011	8	OMT	30.2	1914	1970	1999	22.4	1114
6	Vh012	10	OMT	32.4	1916	1970	1998	23.4	1042
7	Vh013	8	OMT	32.9	1932	1970	1998	18.1	2013
8	Vh014	8	OMT	32.7	1918	1971	1998	23.1	935
9	Vh017	4	OMT	32.9	1936	1971	1985	16.3	2932
10	Vh048	8	OMT	31.0	1934	1977	2001	16.9	2137
11	Vh097	12	OMT	34.0	1955	1981	1994	12.5	2973
12	Ha001	5	OMT	30.2	1938	1965	1998	9.8	3335
13	Pu041	4	OMT	33.0	1934	1964	1999	12.1	1689
14	Pu042	4	OMT	33.0	1924	1964	1999	17.4	1168
15	Nyn1	12	OMT	30.0	1922	1961	1998	14.2	2055
16	Nyn2	4	OMT	34.5	1931	1962	1998	14.7	3247
17	Nyn3	8	OMT	34.7	1926	1962	1999	16.0	2394
18	Nyn4	4	OMT	33.0	1925	1962	1988	15.6	2295
19	Nyn5	4	OMT	33.0	1930	1962	1992	14.3	3402
<i>Syst</i>									
1	102	2	OMT	28.6	1933	1974	1995	15.4	2768
2	107	3	OMT	32.0	1940	1977	1994	16.0	3290

Table 2. Independent variables and the parameter estimates used in the prediction of crown base height (Eq. 7).

Variable	Parameter estimate	Standard error
<i>Fixed effects</i>		
Intercept	-3.2697	0.5169
Tree height (m)	0.4125	0.0865
Stem diameter (cm)	0.3769	0.0841
Height/Diameter	-0.7335	0.0745
Stand density (n ha ⁻¹)	0.1117	0.0539
Stand basal area (m ² ha ⁻¹)	-0.1052	0.2646
Stand volume (m ³ ha ⁻¹)	0.4796	0.1849
Stand age (a)	0.1568	0.1299
<i>Random effects</i>		
Stand	-3.14E-12	0.0143
Plot	-7.07E-13	0.0060

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Table 3. Validation metrics. AMB = absolute model bias (negative values are overestimates, positive underestimates, cm in diameter, m in dominant height, $\text{m}^2 \text{ha}^{-1}$ in basal area, and $\text{m}^3 \text{ha}^{-1}$ in stand volume), RMB = relative model bias, EF = modeling efficiency.

Experiment	Variable	AMB	RMB (%)	EF
<i>Harkas</i>	Mean stem diameter	0.57	2.9	0.95
	Dominant height	0.07	0.3	0.96
	Stand basal area	1.92	5.5	0.79
	Stand volume	29	8.2	0.86
<i>Syst</i>	Mean stem diameter	-0.68	-3.8	0.91
	Dominant height	-0.72	-3.7	0.86
	Stand basal area	-2.62	-7.0	0.86
	Stand volume	20	6.3	0.82

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Supplementary material

S1. Model description

In the TREE module, trees are described in terms of 16 state variables which are bound together through structural rules, such that only a few fully independent variables remain (Table S1). These include tree height, height to the crown base, stem basal area at breast height, and mass of branch heartwood. Temporary deviations of the structural rules may occur due to, e.g, defoliation or abrupt changes in the environment, such as those caused by thinning (Mäkelä 1999). The most important structural rules are the following:

- 1) foliage mass, W_f , is related to crown length, $H - H_C$, through an allometric relationship (Mäkelä and Sievänen 1992):

$$W_f = \xi(H - H_C)^z \quad (S1)$$

where H is tree height, H_C is height to the crown base and ξ and z are parameters.

- 2) sapwood cross-sectional areas of stem (A_s), branches (A_b), and coarse roots (A_c), are related to foliage mass according to the pipe model:

$$W_f = \eta_i A_i \quad (S2)$$

where η_i ($i = s, b, c$) are parameters.

- 3) fine root mass, W_r , is related to foliage mass

$$W_r = \alpha_r W_f \quad (S3)$$

where the parameter α_r depends on site fertility.

- 4) sapwood biomasses are related to sapwood area and mean length, L_i ($i = s, b, c$):

$$W_i = \phi_i A_i L_i \quad (S4)$$

where ϕ_i ($i = s, b, c$) are empirical form factors.

- 5) the mean lengths of the branch and coarse root systems are proportional to crown length and tree height, respectively.

The WHORL module takes in the information from the TREE module and distributes the organ biomasses to whorls on the basis of empirical information about the vertical distributions of biomass in the tree (Mäkelä and Mäkinen 2003; Kantola et al. 2007). Importantly, the foliage biomass is assumed to follow a β -function which moves upward as tree height and crown base rise. This distribution gives rise to sapwood area in whorls, which turns into heartwood as the foliage reduces in the lower whorls. At the same time, stem heartwood accumulates when the wood loses its connection to live foliage.

Detailed stem and branch structure is described in the WHORL module which contains the sapwood and heartwood area and section length as state variables for each whorl. The growth of the whorls is driven by the TREE module, and changes in structure are fed back to the TREE module in the form of aggregated parameter values updated each year (Mäkelä et al. 1997).

The BRANCH module further divides the branch sapwood into individual branches. It computes the number of branches and their size distribution in the new whorls, then keeps track of the sizes of all branches and finally induces branch mortality and shedding. The BRANCH module is fully statistical and has no feedback effect on the rest of the model components.

S2. Parameter estimation

To estimate the parameters of new equations, the model was simulated with a range of plausible values, and a set of parameters providing a qualitatively reasonable output was selected. This was done prior to the quantitative model testing against the PSP data.

The initial values of the parameter ranges were either derived from the literature or set in such a way that the model produced logical responses at the stand level. In general,

different combinations of the parameters could produce virtually the same stand-level responses. As the equations are semi-empirical or phenomenological, the actual values of the parameters are largely unknown and model calibration is needed in order to fix the parameter values. Here, we show the results of stand-level model sensitivity analysis which was used to select the parameter values. These parameters were related to the interactions of tree size classes, and the main influence of the present data was through the initialization procedure. Other model parameters have been reported in previous studies (Kantola et al. 2007; Niinimäki et al. 2012; Mäkelä et al. 2016).

Figure S1 illustrates the Eqn 6 of study. Crown rise occurs if the light level below the crown goes below a threshold, then rapidly accelerates to match height growth as the light levels fall.

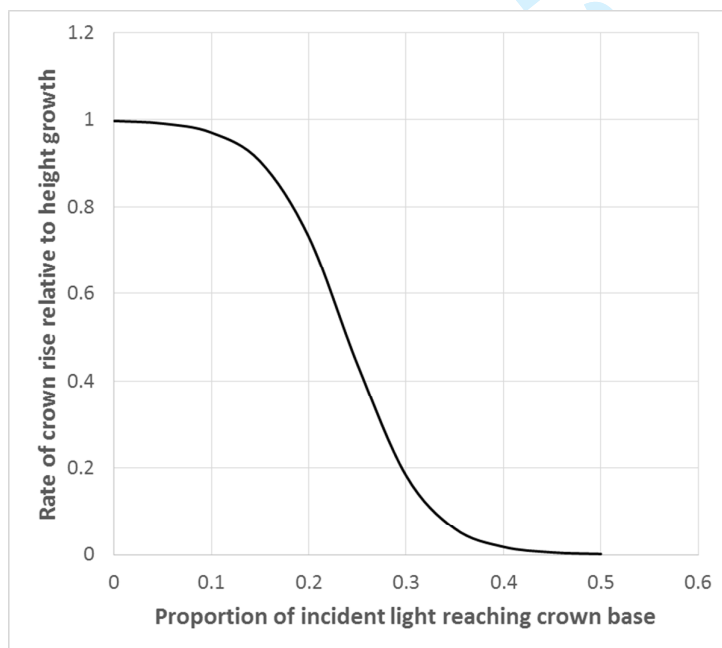


Fig. S1. The vertical axis shows s_C as a function of $f(H_C)$ when $a = 25$ and $f_0 = 0.24$ (default values in simulations).

Table S1. Parameters related to structural acclimation to light

Parameter	Value	Units	Equation
a	25	-	(S3)
f_0	0.24	-	(S3)
ξ_0	0.112	-	(S4)
f_1	0.55	-	(S4)
γ_{b0}	0.20	-	(S5)
A_{tot}	0.7	-	(S5)

The foliage density parameter ξ_0 relates foliage mass to crown length in good and moderate light (Fig. S2), while f_1 is the mean relative light level of the crown that causes the foliage density to decline (Fig. S3). The parameter ξ_0 was allowed to decline in trees in very poor light. Height growth response is very sensitive to the changes of this parameter (Fig. S2).

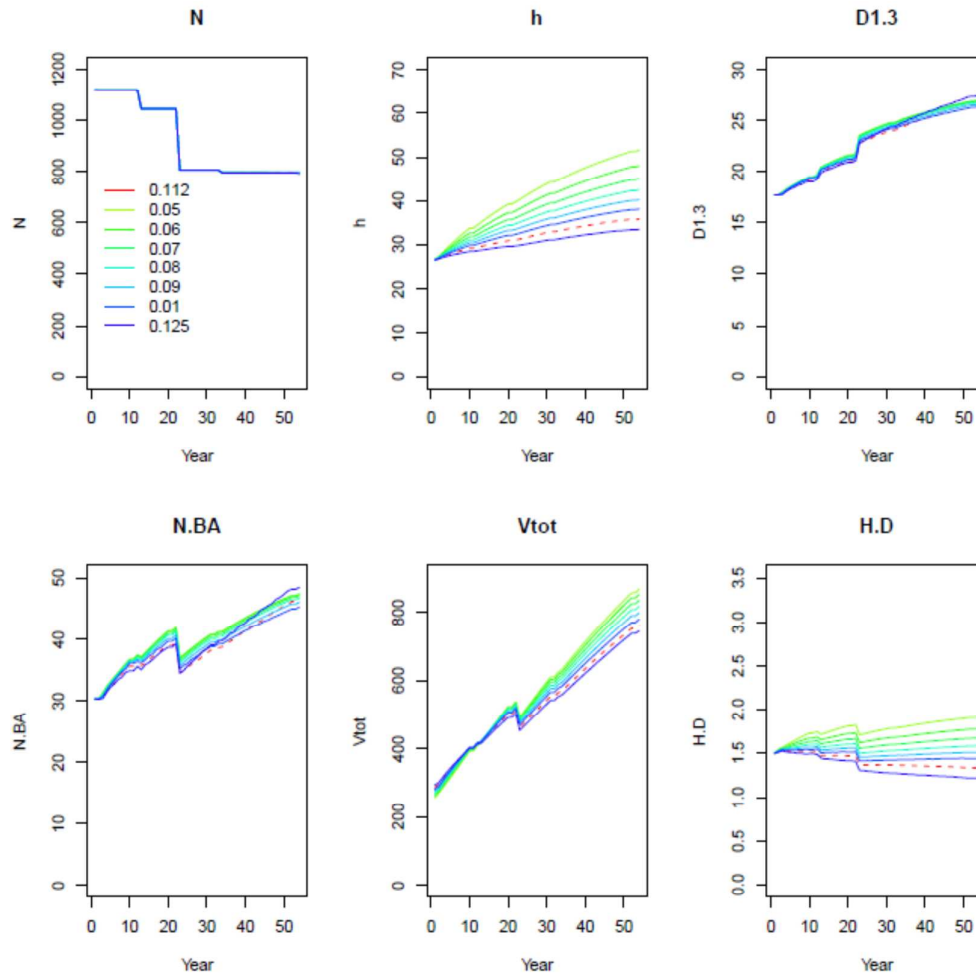


Fig. S2. The effect of ξ parameter to stand variables. Dashed line represents the original value used in simulations. N is stand density (number of trees ha^{-1}), h stand maximum height (m), D1.3 stand mean diameter (cm), N.BA stand basal area ($\text{m}^2 \text{ha}^{-1}$), Vtot stand volume ($\text{m}^3 \text{ha}^{-1}$) and H.D height to diameter ratio.

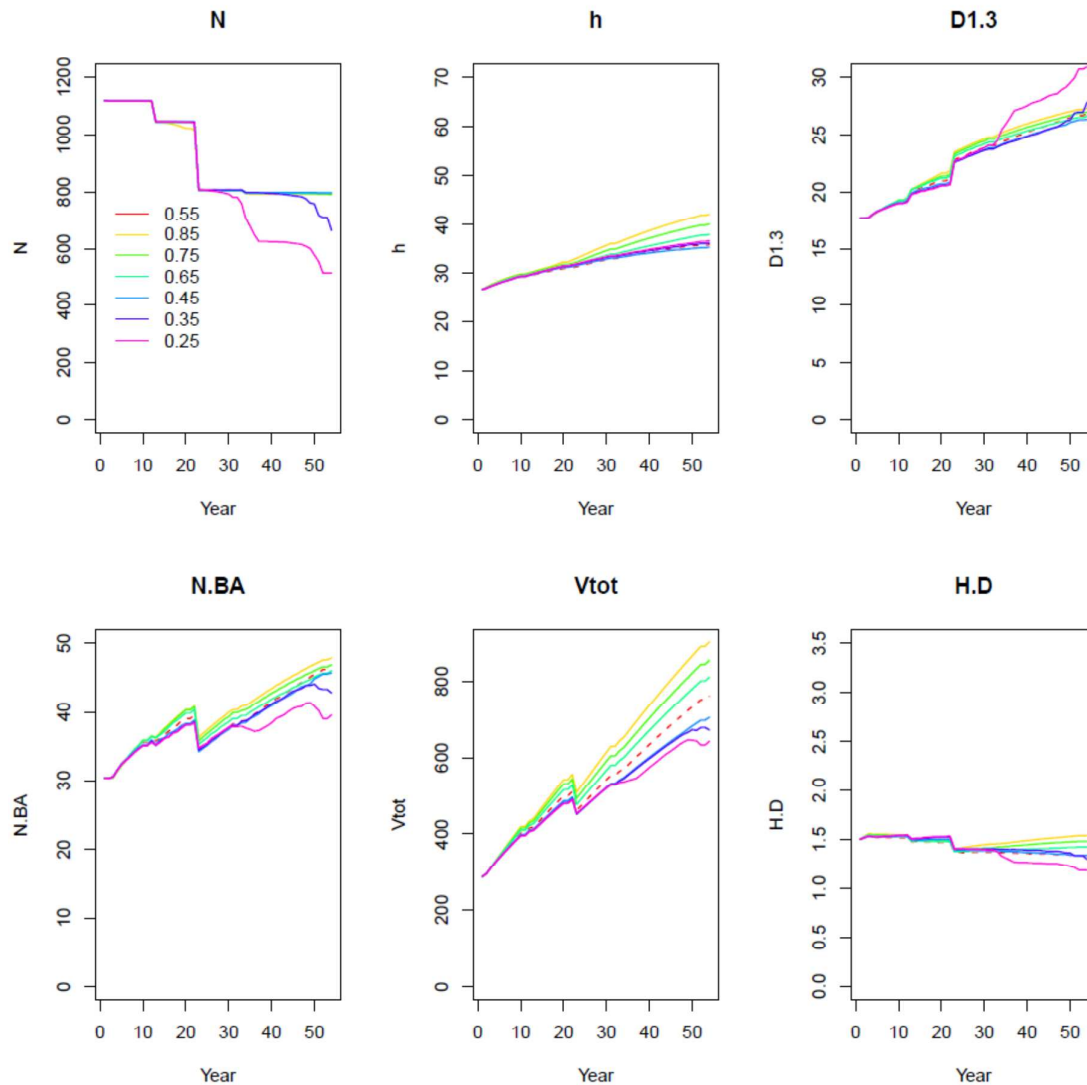


Fig. S3. The effect of f_1 parameter to stand variables. Dashed line represents the original value used in simulations. N is stand density (number of trees ha^{-1}), h stand maximum height (m), $D1.3$ stand mean diameter (cm), $N.BA$ stand basal area ($\text{m}^2 \text{ha}^{-1}$), V_{tot} stand volume ($\text{m}^3 \text{ha}^{-1}$) and $H.D$ height to diameter ratio.

In the model, crown rise follows the height growth as shown in Eqns S2 and S3 and Figure S1. The parameter f_0 in s_C function defines the light level at which the rate of crown rise is half of the height growth rate (Fig. S4) while the parameter a determines the steepness of the curve: if a is large, the switch from no crown rise to maximum crown rise is abrupt, whereas for small a the effect of declining light is gradual (Fig. S5).

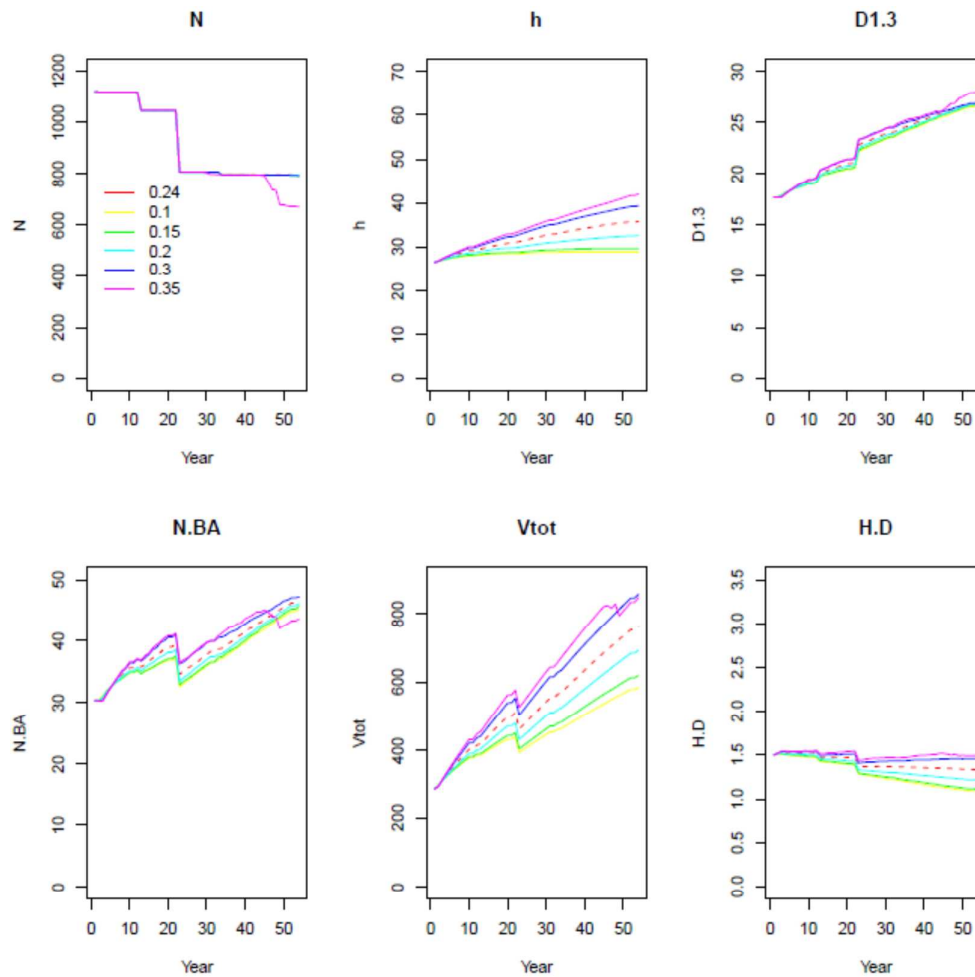


Fig. S4. The effect of f_0 parameter to stand variables. Dashed line represents the original value used in simulations. N is stand density (number of trees ha^{-1}), h stand maximum height (m), D1.3 stand mean diameter (cm), N.BA stand basal area ($\text{m}^2 \text{ha}^{-1}$), Vtot stand volume ($\text{m}^3 \text{ha}^{-1}$) and H.D height to diameter ratio.

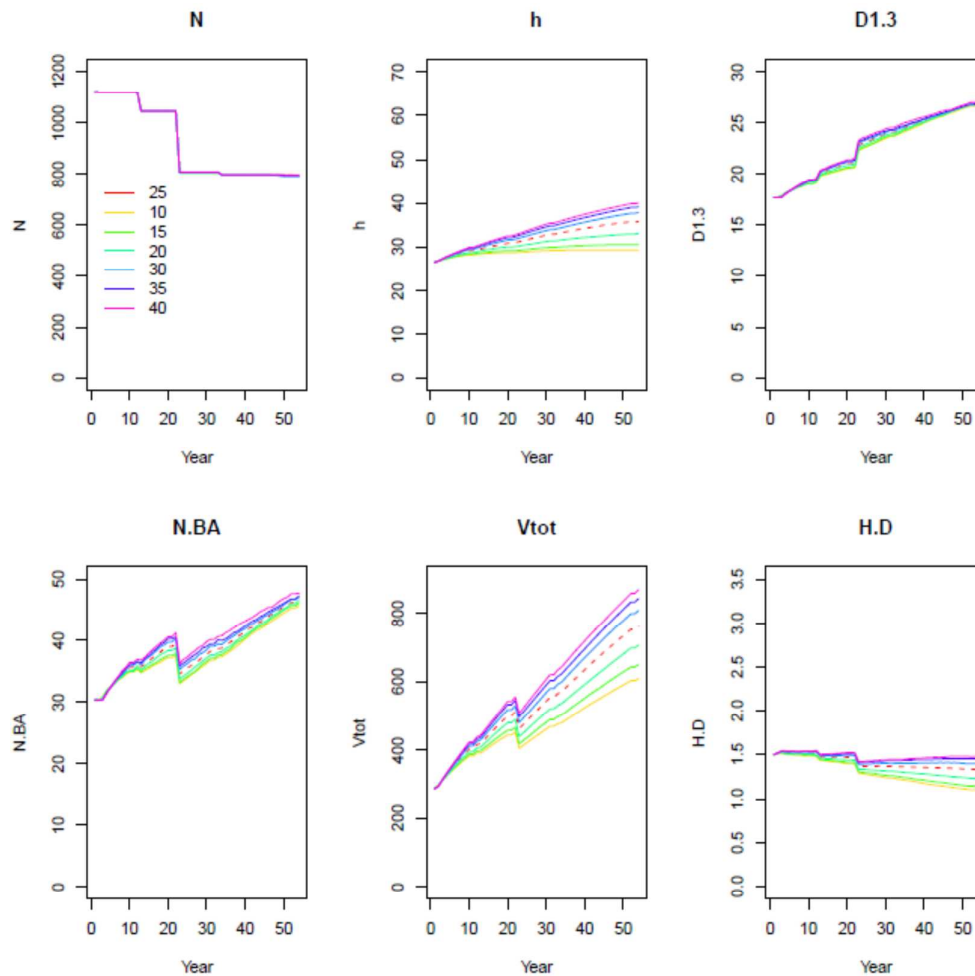


Fig. S5. The effect of α parameter to stand variables. Dashed line represents the original value used in simulations. N is stand density (number of trees ha^{-1}), h stand maximum height (m), D1.3 stand mean diameter (cm), N.BA stand basal area ($\text{m}^2 \text{ha}^{-1}$), Vtot stand volume ($\text{m}^3 \text{ha}^{-1}$) and H.D height to diameter ratio.

The growth of branch length was assumed to be regulated by the stand crown coverage A_{tot} . Parameter $A_{\text{tot},0}$ describes the crown ratio above which average branch length relative to crown length starts to decline. As shown in Fig. S6 the model response to $A_{\text{tot},0}$ is conservative especially in height growth. Only very low values cause drastic changes in model response.

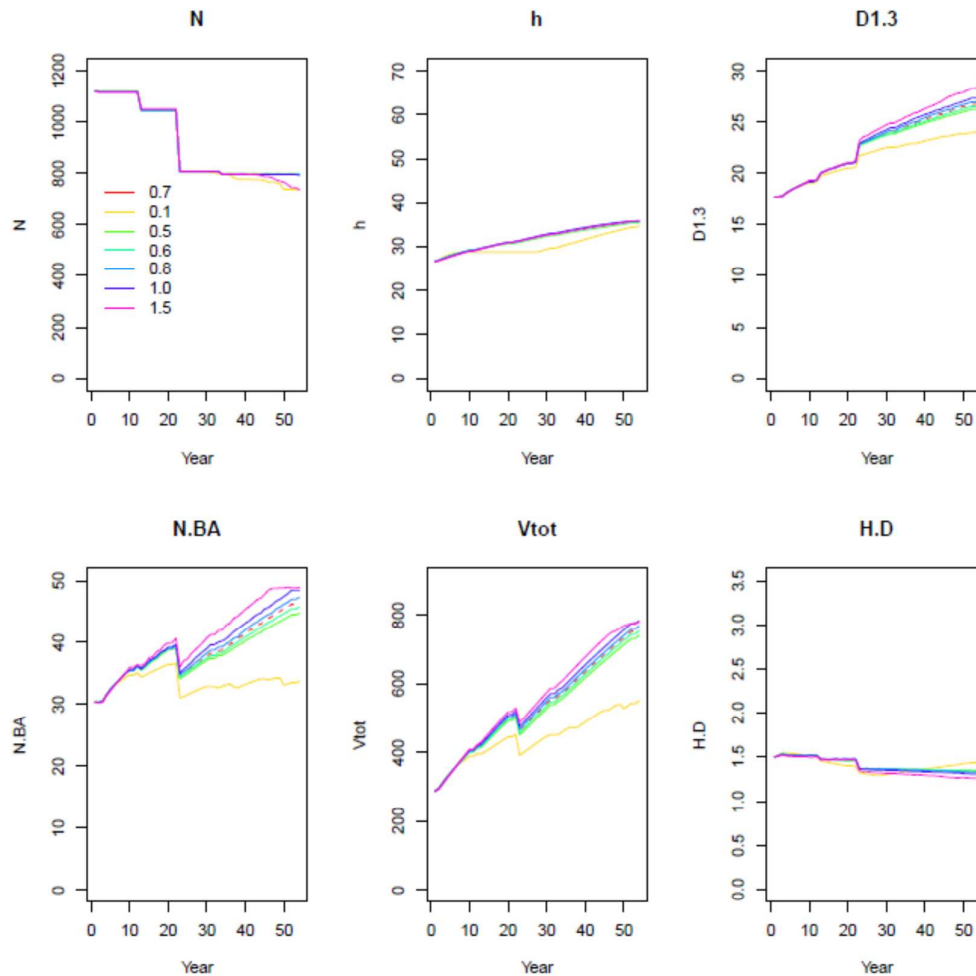


Fig. S5. The effect of $A_{tot,0}$ parameter to stand variables. Dashed line represents the original value used in simulations. N is stand density (number of trees ha^{-1}), h stand maximum height (m), D1.3 stand mean diameter (cm), N.BA stand basal area ($m^2 ha^{-1}$), Vtot stand volume ($m^3 ha^{-1}$) and H.D height to diameter ratio.