

1 **Phenotypical and molecular changes induced by carbamazepine and propranolol on**  
2 **larval stages of *Mytilus galloprovincialis***

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20

21 **Abstract**

22

23 The possible impact of carbamazepine (CBZ) and propranolol (PROP), two widespread  
24 pharmaceuticals in the aquatic environment, were investigated on morphology and gene  
25 transcription of early larvae of *Mytilus galloprovincialis*. Pharmaceuticals were first tested  
26 in a wide concentration range (from 0.01 to 1000 µg/L) through the 48-hpf embryotoxicity  
27 assay. The results showed that both compounds significantly affected embryo  
28 development from environmental concentrations. Although similar EC<sub>50</sub> were obtained, (≅  
29 1 µg/L) CBZ induced a progressive increase in embryo malformations, whereas PROP  
30 apparently showed greater impacts in terms of arrested development and embryo mortality  
31 at higher concentrations (>10 µg/L). Transcriptional analyses of 17 genes involved in  
32 different physiological functions in mussels and/or in their response to environmental  
33 contaminants, were performed at 24 and 48 h pf at two selected concentrations of CBZ  
34 and PROP (0.01 and 1 µg/L). Both compounds induced down-regulation of shell-specific  
35 and neuroendocrine related transcripts, while distinct effects were observed on  
36 antioxidant, lysosomal, and immune-related transcripts, also depending on the larval stage  
37 investigated. The results demonstrate that CBZ and PROP can affect development and  
38 gene transcription in mussel early larvae at environmental concentrations.

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40 **Keywords:** mussel embryo development; pharmaceutical; carbamazepine; propranolol;  
41 embryotoxicity; gene transcription.

## 42 1 Introduction

43        Pharmaceuticals are biologically active molecules used in human and veterinary  
44 medicine, which are partially excreted as parent compounds or active metabolites and  
45 enter wastewater treatment plants (WWTP). Given their increased consumption and low-  
46 retention in WWTP, as well as the development of more powerful analytical techniques,  
47 pharmaceuticals have been widely detected in the marine environment, mainly in coastal  
48 areas (Fabbri and Franzellitti, 2016). Although half-life of these compounds is relatively  
49 short, the marine biota is exposed to a continuous input, raising concern on the potential  
50 biological impacts (Fent et al., 2006). Research efforts over the last decade also pointed  
51 out that pharmaceuticals may affect marine fauna at environmental concentrations when  
52 specific responses related to the therapeutic targets or the modes of action known in  
53 humans are investigated (Brooks, 2018; Fabbri and Franzellitti, 2016).

54        Amongst the pharmaceuticals routinely detected in coastal environments,  
55 carbamazepine (CBZ) and propranolol (PROP) attracted special attention when it was  
56 proved that they induced sublethal effects in marine invertebrates, including mussels  
57 (Fabbri and Franzellitti, 2016), and a link between these effects and the therapeutic modes  
58 of action in humans was revealed (Franzellitti et al., 2013, 2011; Martin-Diaz et al., 2009).

59        Carbamazepine (CBZ) is an antiepileptic drug which poorly retained during wastewater  
60 treatments (below 10%), thus it is found in all aquatic compartments, including drinking  
61 waters, at the low  $\mu\text{g/L}$  concentration range (Bahlmann et al., 2012; Calisto et al., 2011;  
62 Metcalfe et al., 2014). In seawater, CBZ was documented in harbours (up to 321 ng/L),  
63 coastal waters (up to 1400 ng/L) and sediments (up to 89 ng/g), and in river estuaries (up  
64 to 178 ng/L) (Fabbri and Franzellitti, 2016). Propranolol (PROP) is administered to  
65 counteract hypertension and other cardiovascular pathologies. It is found in surface waters  
66 (ng/L range), in transitional waters (up to 142 ng/L), coastal waters (up to 6329 ng/L) and

67 sediments (up to 0.9 ng/g) (Ashton et al., 2004; Fabbri and Franzellitti, 2016; Wille et al.,  
68 2011). Data on occurrence of both pharmaceuticals in seawater are summarized in Table  
69 S1, Supplemental material.

70 Recent studies indicated that embryos/larvae from different species can be utilized as  
71 promising experimental models to establish the sensitivity of marine species towards  
72 pharmaceuticals (Balbi et al., 2016; Di Poi et al., 2017, 2014; Estévez-Calvar et al., 2017;  
73 Fabbri et al., 2014; Franzellitti et al., 2017; Minguez et al., 2014). Compounds with widely  
74 different therapeutic effects and chemical reactivity have been shown to affect early larval  
75 development in the Mediterranean mussel, *Mytilus galloprovincialis* (Balbi et al., 2016;  
76 Estévez-Calvar et al., 2017; Fabbri et al., 2014). In particular, 17 $\beta$ -estradiol and diclofenac  
77 altered the transcription of genes involved in shell biogenesis and serotonin signaling  
78 (Balbi et al., 2018, 2016).

79 This study is addressed to the potential effects of CBZ and PROP on early larval  
80 development of *M. galloprovincialis*. Phenotypical outcomes of the pharmaceuticals were  
81 investigated in a wide concentration range (0.01 – 1000  $\mu$ g/L) through the standardized  
82 48-hpf embryotoxicity assay (Fabbri et al., 2014) which evaluates the impairment of normal  
83 D-larvae development. Moreover, transcriptional analyses of 17 gene products at different  
84 developmental stages (24 and 48 h post fertilization, pf) (Balbi et al., 2017a, 2016) were  
85 carried out after exposure to two selected CBZ or PROP concentrations (0.01 and 1  $\mu$ g/L).  
86 As both compounds may affect physiological pathways in bivalves (Fabbri and Capuzzo,  
87 2010; Franzellitti et al., 2017; Franzellitti and Fabbri, 2013), this latter experimental  
88 approach aimed to disclose early signs of CBZ/PROP interactions with developmental  
89 processes of mussels.

90

## 91 2 Methods

### 92 2.1 Chemicals

93 Carbamazepine (CBZ), and ( $\pm$ )-propranolol hydrochloride (PROP) were at the  
94 molecular biology grade (> 99 % purity) and were purchased from Sigma Aldrich (Milan,  
95 Italy). The DirectZol kit was from Zymo Research (Freiburg, Germany). The Qubit RNA  
96 assay and the Qubit protein assay were from Thermo Fisher (Milan, Italy). The iScript  
97 supermix and iTaq Universal master mix with ROX were from Biorad Laboratories (Milan,  
98 Italy). The Tri-Reagent, dimethylsulfoxide (DMSO), and any other reagent was from Sigma  
99 Aldrich (Milan, Italy).

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### 101 2.2 Mussels and gamete collection

102 Sexually mature mussels (*M. galloprovincialis* Lam.) were obtained from a certified  
103 mussel farm (Cooperativa Copr.al.mo, Cesenatico, Italy) and acclimatized in static tanks  
104 containing aerated artificial 35-psu seawater (ASTM, 2004) at  $16 \pm 1$  °C (1 L/animal).  
105 Procedures for gamete collection, oocyte fertilization and larvae handling were according  
106 to Fabbri et al. (2014). Fertilizations were performed with an egg:sperm ratio 1:10 in 96-  
107 well (embryotoxicity assay; 200  $\mu$ L final volume) or 6-well (transcriptional analyses; 10 mL  
108 final volume) cell culture plates. After 30 min, fertilization success (n. fertilized eggs / n.  
109 total eggs  $\times$  100) was verified by microscopical observation (>85%) and carbamazepine  
110 (CBZ) or propranolol (PROP) were added to fertilized eggs in the proper wells from  
111 concentrated stock solutions prepared in DMSO (CBZ) or sterile milli-Q water (PROP),  
112 considering solubility limits of the compounds in the selected medium, and diluted in ASW  
113 to reach the final nominal concentrations tested. Concentrations of the stock solutions  
114 were verified by LC-MS/MS (Castiglioni et al., 2005). They were within a 8% discrepancy  
115 compared to the nominal values. DMSO final concentration never exceeded 0.1% v/v and

116 it did not significantly affect the biological endpoints analysed (data not shown). Control  
117 (ASW) and solvent (ASW+DMSO; 0.1% v/v) exposed samples were run in parallel. After  
118 pharmaceutical addition, samples were maintained at  $16 \pm 1$  °C for 48 h with a 16 h:8 h  
119 light:dark photoperiod according to ASTM (2004).

### 120 2.3 *Embryotoxicity assay*

121 The 48-h embryotoxicity assay (ASTM, 2004) was performed according to Fabbri et  
122 al. (2014) and using a wide range of nominal concentrations from 0.01 to 1000 µg/L. Four  
123 independent experiments were performed, each consisting of 6 technical replicates (N =  
124 4). At the end of the incubation time, samples were fixed with buffered formalin (4%). For  
125 each well, all larvae were examined by optical and/or phase contrast microscopy using an  
126 inverted microscope (Olympus IX53, Olympus, Milan, Italy) at 40X magnification.  
127 Observations were carried out by an operator blind to the experimental conditions.  
128 Classification of larvae morphotypes was performed according to Balbi et al. (2017a) and  
129 Fabbri et al. (2014), with the recorded endpoint being the percentage of normal D-shape  
130 larvae in each well over the total. The acceptability of assay results was achieved with  
131 controls showing a > 75% normal D-shape larvae (ASTM, 2004).

132 According to the ASTM guidelines (ASTM, 2004), embryo viability at 48 h pf was  
133 assessed as the percentage of live larvae with misshapen or otherwise malformed shells  
134 and completely/normally developed shells. Empty shells, larvae with incompletely  
135 developed shells or larvae retained at the trocophora stage were considered non-viable  
136 because retarded development is likely to reduce survival in the natural environment  
137 (ASTM, 2004).

138

## 139 2.4 RNA extraction and qPCR analyses

140 Treatments for mRNA expression analyses were carried out at 0.01 µg/L and 1 µg/L.  
141 Four independent experiments were performed, each consisting of 3 technical replicates  
142 (N = 4). Total RNA extraction from embryos at 24 h and 48 h pf was performed according  
143 to Balbi et al. (2016). RNA concentration and quality were verified using the Qubit RNA  
144 assay through the Qubit 2.0 system (Thermo Fisher, Milan, Italy) and electrophoresis  
145 using a 1.2% agarose gel under denaturing conditions. First strand cDNA for each sample  
146 was synthesized from 1 µg total RNA using the iScript supermix following manufacturer's  
147 instructions. Target transcripts are listed in Table S2, Supplemental material. Primer pairs  
148 and protocols employed for quantitative real time PCR (qPCR) assays are reported in  
149 previous studies (Balbi et al., 2017a, 2016; Capolupo et al., 2018). A preliminary stability  
150 analysis performed according to Balbi et al. (2016) selected the pair *helicase/elongation*  
151 *factor 1α* as the reference gene products for qPCR data normalization. A comparative C<sub>T</sub>  
152 method was used for fold-change calculations (Schmittgen and Livak, 2008) through the  
153 StepOne software tool (Thermo Fisher, Milan, Italy). Data are reported as mean ± SEM of  
154 log<sub>2</sub>-transformed fold changes with respect to controls within each post-fertilization time.

155 Expressions of transcripts related to antioxidant responses, immune responses and  
156 lysosomal functions were also evaluated in unfertilized eggs as well as in 24- and 48-h pf  
157 embryos grown under normal physiological conditions, to account for their baseline  
158 regulation across early embryo development (Fig. S1, Supplemental material)

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## 160 2.5 Statistics

161 Data from embryotoxicity and embryo viability were analysed using the non-parametric  
162 1-way ANOVA (Kruskal-Wallis test) and the Mann-Whitney U-test for pairwise  
163 comparisons by the GraphPad Prism 6 software (GraphPad Inc.). Non-accomplishment of

164 parametric ANOVA assumptions was verified (Normality: Shapiro-Wilk's test; equal  
165 variance: F-test). EC50 values for the embryotoxicity were calculated using the Hill  
166 equation for non-linear regression through the Excel™ macro REGTOX (Vindimian, 2012).

167 qPCR data were analysed with the REST software (Pfaffl et al., 2002) that employs  
168 randomisation tests with pairwise reallocations to assess the significance differently  
169 expressed transcripts between each treatment-exposed group and the controls. Further  
170 comparisons between pair of treatments were performed using the Mann-Whitney U-test.  
171 In all approaches,  $p < 0.05$  was set as the threshold level of statistical significance.

172 Data from qPCR investigations were further submitted to permutation multivariate  
173 analysis of variance (PERMANOVA) and distance-based redundancy linear modeling  
174 (DISTLM) as detailed by (Balbi et al., 2016). Factors considered were “developmental  
175 stage” and “treatment” (CBZ or PROP). DISTLM with a test of marginality was also  
176 performed to account for the contributions of the functional groups of transcripts listed in  
177 Table S2 in explaining the total variance observed in the CBZ/PROP dataset (Rasika  
178 Wathsala et al., 2018).

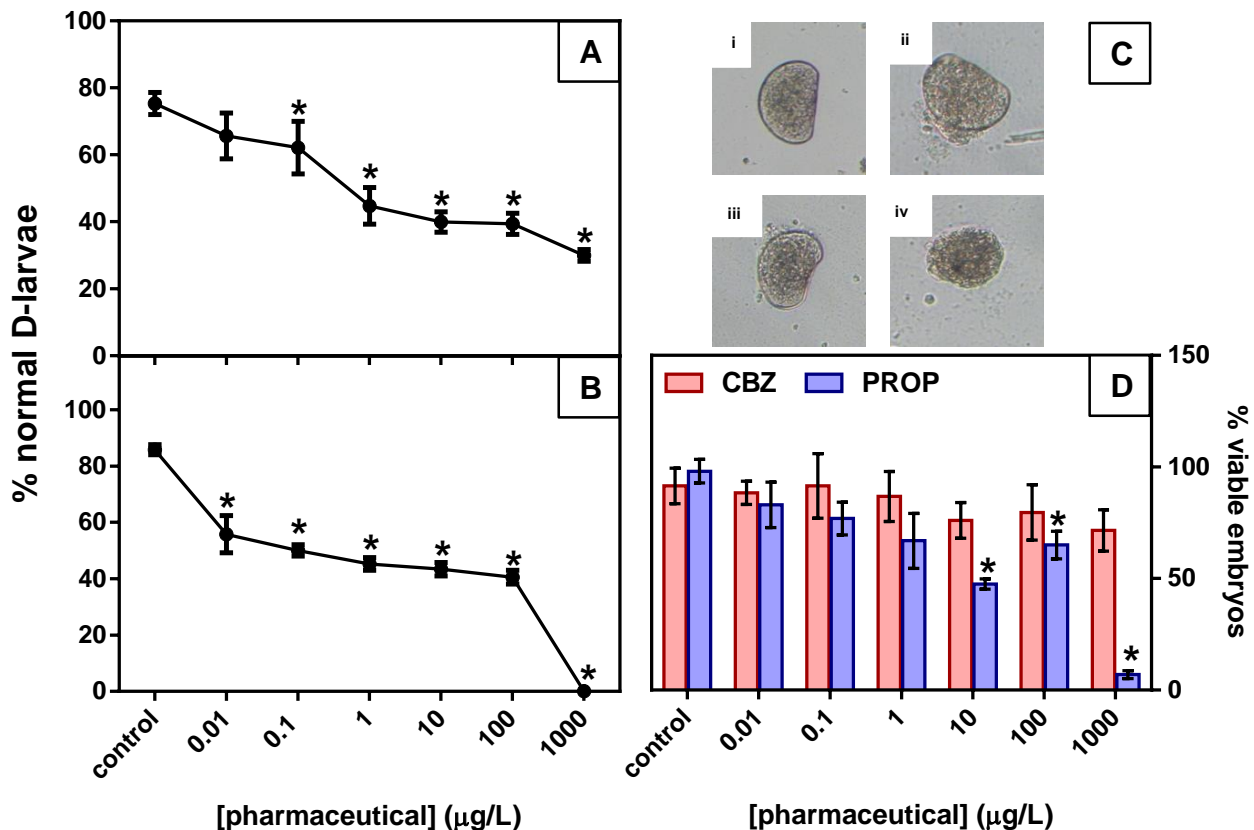
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## 180 **3 Results**

### 181 *3.1 Effects of CBZ and PROP on embryo development*

182 Fertilized eggs were exposed to different concentrations (from 0.01 µg/L to 1000 µg/L) of  
183 PROP or CBZ in 96-microwell plates, and the percentage of normal D-larvae was  
184 evaluated 48 hpf (Fig. 1). Both compounds induced a significant decrease in percentage of  
185 normally developed embryos from 0.1 µg/L (CBZ) and 0.01 µg/L (PROP) ( $p < 0.05$ ), with  
186 values decreasing thereafter (Fig. 1A, B).





187

188 **Fig. 1. Effects of different concentrations of CBZ or PROP (0.01 – 1000 µg/L) on *M. galloprovincialis***  
 189 **embryo development.** (A, CBZ; B, PROP) Results of the 48-h embryotoxicity assay. (C) Representative  
 190 light microscope images (32X magnification) showing the main types of abnormalities observed: (i) normal  
 191 48 h pf D-shaped veliger; (ii) D-shaped 48 h pf veliger with protruding mantle; (iii) D-shaped veliger with shell  
 192 hinge malformations; (iv) embryo arrested at the trochophora stage (characteristic of PROP-treated  
 193 samples). (D) Percentage of embryo viability at 48 h pf. All data are reported as mean ± SEM of 4  
 194 experiments carried out in 96-multiwell plates (6 replicate wells for each sample) (N = 4). \*p < 0.05 vs control  
 195 (Mann-Whitney U-test). **Colored figure is intended only for the online and PDF version.**  
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197 Although similar EC50 values were obtained for CBZ and PROP ( $0.82 \pm 0.36$  and  $1.34 \pm$   
 198  $0.43$  µg/L, respectively), CBZ mainly induced the development of malformed D-veligers at  
 199 all the concentrations tested, whereas exposure to PROP resulted in a variable proportion  
 200 of malformed/immature embryos at different concentrations, with all embryos retained at  
 201 the trochophora stage at the highest concentration tested (1000 µg/L) (Fig. 1C).  
 202 Furthermore, while no significant mortality due to CBZ was observed, the average

203 proportion of viable embryos at 48 h pf was significantly reduced at 10-1000 µg/L PROP  
204 (Fig. 1D).

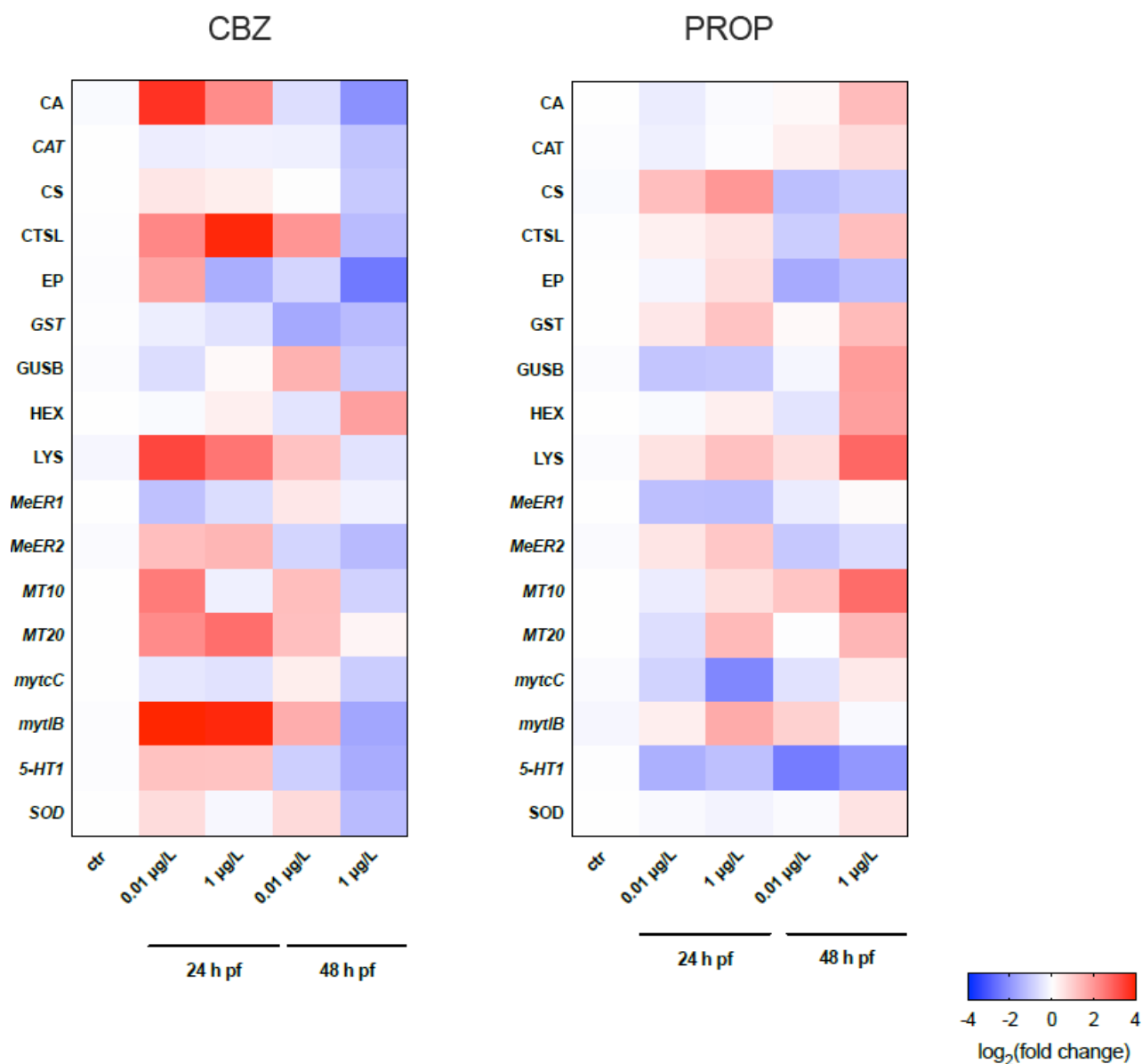
### 205 3.2 *Effects of CBZ and PROP on gene transcription*

206 Gene transcription was evaluated in samples exposed to CBZ or PROP at 0.01 and 1  
207 µg/L. The former concentration encompasses the environmental levels of both compounds  
208 (10 ng/L, Table S1); the latter is close similar to the EC50 values obtained for both  
209 compounds in the embryotoxicity assay, and also represents the highest range of  
210 concentrations measured in coastal environments (1000 ng/L, Table S1).

211 Selected transcripts related to different physiological functions were evaluated:  
212 *Antioxidant response*: Catalase, *CAT*; Glutathione transferase-pi, *GST*; Metallothionein 10  
213 and 20, *MT10* and *MT20*; Superoxide dismutase, *SOD*. *Immune response*: Lysozyme,  
214 *LYS*; Mytilin B, *MytIB*; Myticin C; *MytcC*. *Lysosomal response* Cathepsin L, *CTSL*; β-  
215 Glucuronidase, *GUSB*; Hexosaminidase, *HEX*. *Putative Neuroendocrine signaling*: Type 1  
216 estrogen receptor, *MeER1*; Type 2 estrogen receptor, *MeER2*; Type 1 serotonin (5-HT)  
217 receptor, *5-HT1*. *Shell biogenesis*: Extrapallial protein, *EP*; Carbonic anhydrase, *CA*.  
218 (Balbi et al., 2018, 2017a, 2016) (Details are provided in Table S2).

219 To help comparing overall transcriptional responses among transcripts as well as datasets,  
220 fold change variations (log<sub>2</sub>-transformed) were subjected to similarity analysis which  
221 generated heatmaps describing the overall transcriptional responses to CBZ and PROP  
222 (Fig. 2). Details of results and statistics are reported in Fig. S2.

223 Exposure to CBZ for 24h significantly up-regulated a series of transcripts (*CA*, *CTSL*, *EP*,  
224 *LYS*, *HEX*, *MeER2* *Mt10*, *Mt20*, *MytB*, *5HT1* and *SOD*) at 0.01 µg/L. The effects of 0.01  
225 µg/L CBZ on antioxidant related transcripts were slightly reduced at 48 h pf, except for  
226 *GST* which appeared significantly downregulated.



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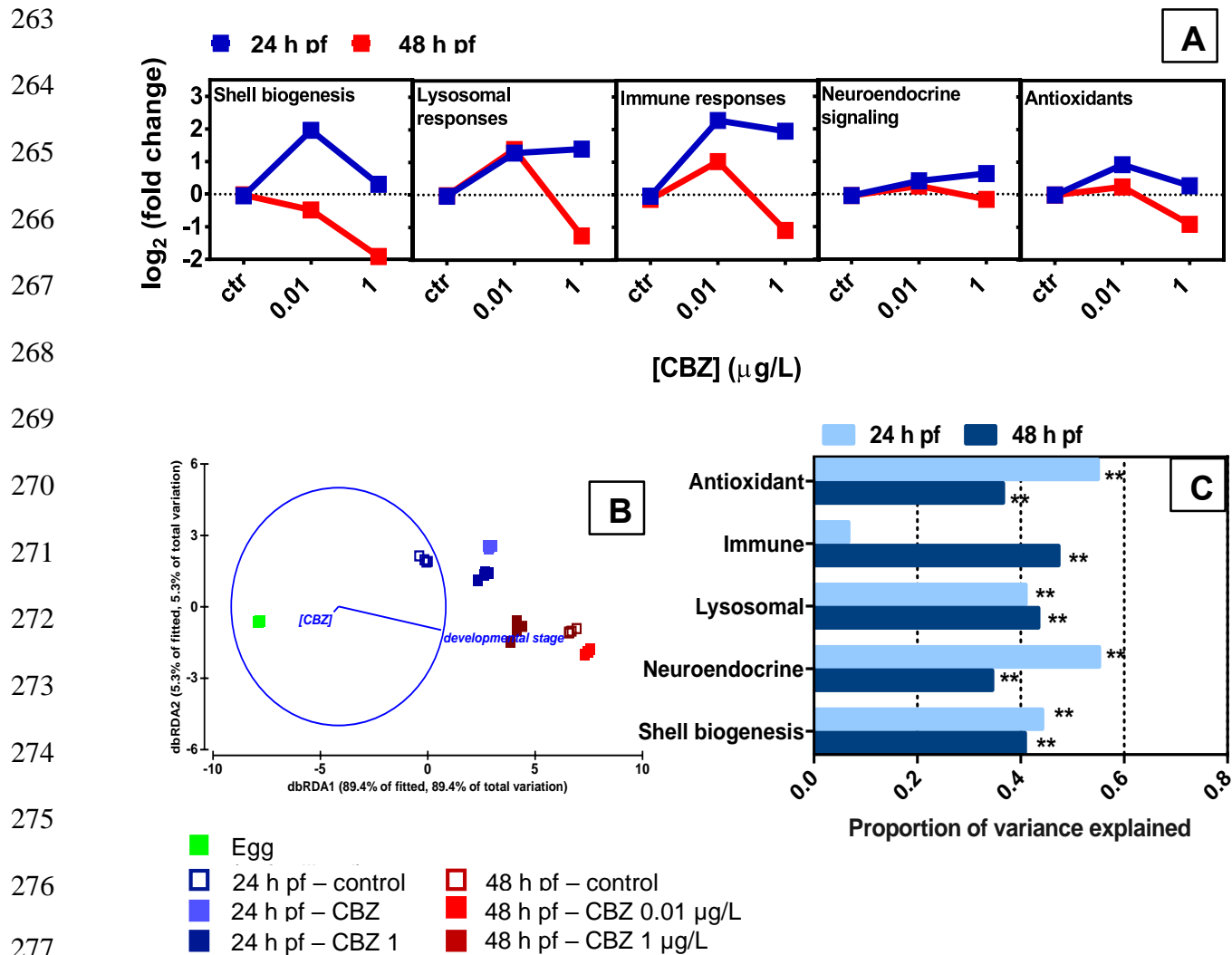
228 **Fig. 2** Heatmaps comparing changes in mRNA levels in early larval stages of *M. galloprovincialis* after  
 229 exposure to CBZ or PROP. Clustering was performed using fold change variations (log<sub>2</sub>-transformed) of  
 230 target transcripts compared to unfertilized eggs (ctr). Colors represent relative expression levels with respect  
 231 to controls. Transcripts are listed in alphabetical order (see Fig. S2 for detailed data and statistics).  
 232

233 Significant upregulation was observed for CTSL and GUSB, and downregulation for EP,  
 234 MeER2, and 5HT1 (Fig. 2, Fig. S2). Significant up regulation of CA, CTSL, LYS, MT20 and  
 235 MytB transcripts was observed after 24 h pf at 1 µg/L CBZ. A general downregulation is  
 236 shown at 48 h pf, with the exception of HEX (up-regulated) and MeER1 and MT20  
 237 (unchanged).

238 Exposure to PROP for 24h differently affected gene transcription, with significant down  
239 regulation of GUSB, MeER1 and 5HT1, and up-regulation of CS transcripts at 0.01 µg/L  
240 (Fig. 2, Fig. S2). The effects of 0.01 µg/L PROP were generally reduced at 48 h pf, except  
241 for *GST* which appeared significantly down regulated (Fig. 2, Fig. S2). The down  
242 regulation of 5HT1 transcript was further enhanced at 48 h pf, while MeER2 and CS  
243 became significantly down-regulated (Fig. 2, Fig. S2). After exposure to 1 µg/L PROP, at  
244 24 h pf significant upregulation was observed for CS, EP, GST, LYS, MT10, MT20, Myt1B,  
245 and MeER2, and down regulation for GUSB, MeER1, Myt1C, and 5HT (Fig. 2, Fig. S2). At  
246 48h pf, 1 µg/L PROP caused significant up regulation of all transcripts related to the  
247 antioxidant response (CAT, GST, SOD, Mt10 and MT20), lysosomal function (CTSL,  
248 GUSB and HEX), and of CA and LYS; while a significant downregulation is shown for  
249 5HT1, CS and EP (Fig. 2, Fig. S2).

250 PERMANOVA and permutation t-test analyses performed on qPCR data showed  
251 that effects of either CBZ or PROP were significant in both embryo stages and at both  
252 concentrations tested ( $p < 0.05$ ; Table S3 and Table S4, Supplemental material). A  
253 significant interaction between the two factors (developmental stage and CBZ/PROP  
254 treatment) was also reported ( $p < 0.05$ ; Table S3, Supplemental material).

255 Fold change variations ( $\log_2$  transformed) of target transcripts were averaged by  
256 functional group, as defined in Table S2, and the resulting concentration-related  
257 transcriptional profiles at each developmental stage reported in Fig. 3A (CBZ) and Fig. 4A  
258 (PROP). CBZ treatments resulted in complex transcriptional profiles, with consistent up-  
259 regulations at 24 h pf (lysosomal responses, immune responses, and putative  
260 neuroendocrine signaling), down-regulations at 48 h pf (shell biogenesis), and bimodal  
261 effects in both embryo stages depending on the tested concentrations (24 h pf: shell  
262 biogenesis; 48 h pf: lysosomal responses, immune responses)



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279 **Fig. 3. Overview of the transcriptional response to CBZ across mussel embryo development.** (A)  
280 Graphs report the mRNA expression changes (full squares representing mean values) and the average trend  
281 of variations (lines) within each group of transcripts (defined in Table S2). Detailed data for transcript  
282 expression changes and related statistics are reported in Fig. S2 (B) Distance-based redundancy (DISTLM)  
283 modeling with distance-based redundancy analysis (dbRDA) exploring the amount of the variation in gene  
284 transcription to be attributed to CBZ treatment of *M. galloprovincialis* embryos at different developmental  
285 stages (Euclidean Distance resemblance matrix, 999 permutations). (C) DISTLM analysis showing  
286 contribution of each functional group to the total variance observed in the CBZ dataset. DISTLM used the  
287 BEST selection procedure and adjusted R<sup>2</sup> criteria. Asterisks indicate level of statistical significance related  
288 to the result (\*\*p < 0.01; \*p < 0.05). **Colored figure is intended only for the online and PDF version**  
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290 Distance-based linear model (DISTLM) analyses revealed that expression profiles were  
291 strongly dependent on the time post-fertilization that explained about 89.4% total variation  
292 (Fig. 3B). Nevertheless, CBZ treatment accounted for about 5% total variation explaining  
293 the observed transcriptional changes at the 24 h and 48 h pf (Fig. 3B). DISTLM analysis  
294 by functional groups also revealed that expression patterns of transcripts involved in

295 antioxidant responses, lysosomal responses, putative neuroendocrine signaling, and shell  
 296 biogenesis explained the observed effects of CBZ at 24 h and 48 h pf (Fig. 3C).  
 297 Responses to CBZ of immune related gene products reached relevance at 48 h pf ( $p <$   
 298 0.05) (Fig. 3C).

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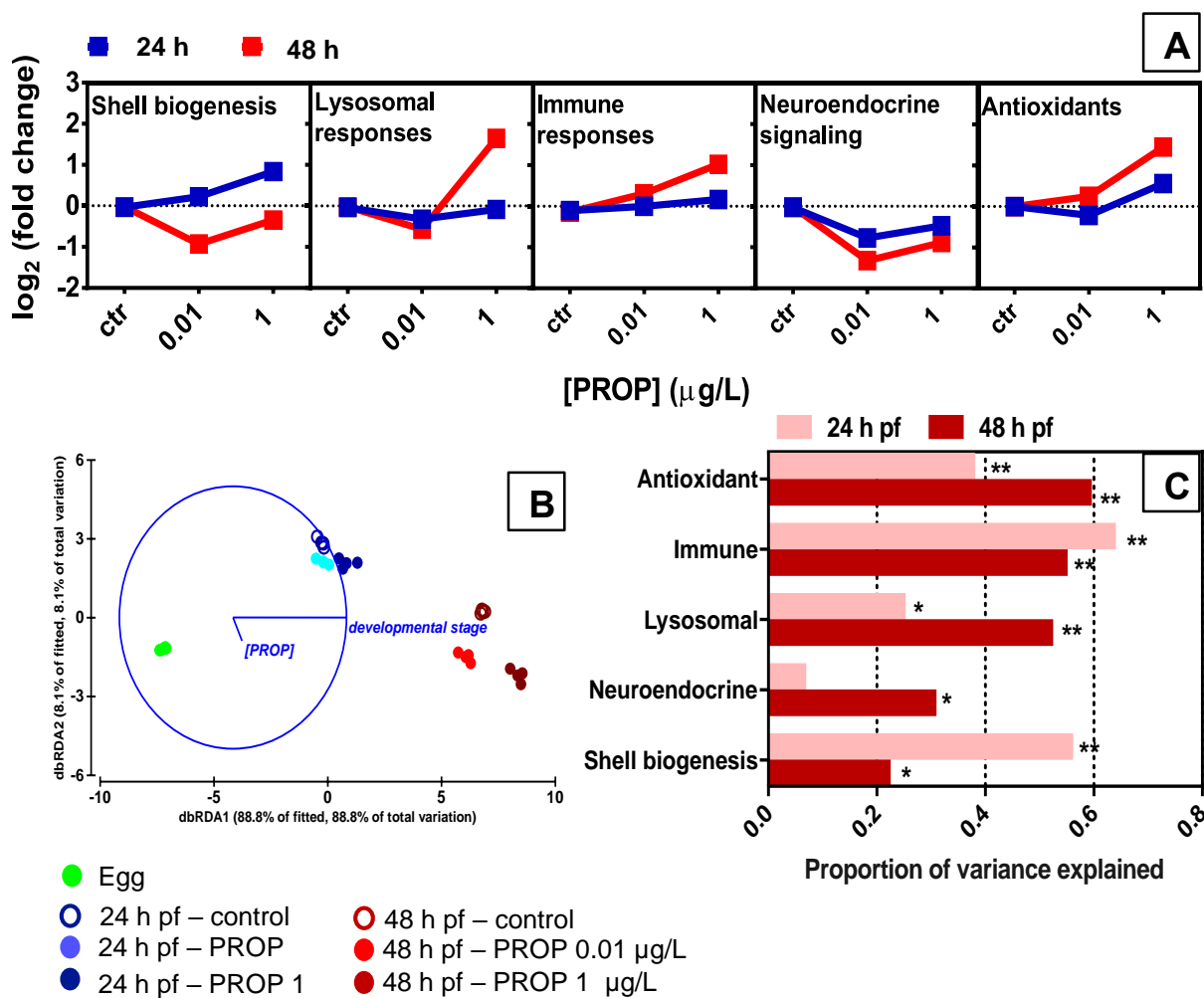
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321 **Fig. 4. Overview of the transcriptional response to PROP across mussel embryo development.** (A)  
 322 Graphs report the mRNA expression changes (full squares representing mean values) and the average trend  
 323 of variations (lines) within each functional group (defined in Table S2). Detailed data for transcript expression  
 324 changes and related statistics are reported in Fig. S2 (B) Distance-based redundancy (DISTLM) modeling  
 325 with distance-based redundancy analysis (dbRDA) exploring the amount of the variation in gene transcription  
 326 to be attributed to PROP treatment of *M. galloprovincialis* embryos at different developmental stages  
 327 (Euclidean Distance resemblance matrix, 999 permutations). (C) DISTLM analysis showing contribution of  
 328 each functional group to the total variance observed in the PROP dataset. DISTLM used the BEST selection  
 329 procedure and adjusted R<sup>2</sup> criteria. Asterisks indicate level of statistical significance related to the result (\*\*p  
 330 < 0.01; \*p < 0.05). **Colored figure is intended only for the online and PDF version**

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332 PROP treatments at 24 h pf caused overall up-regulation of transcripts involved in  
333 antioxidant responses and shell biogenesis, down-regulation of transcripts involved in  
334 putative neuroendocrine signaling, and globally unchanged expression patterns of  
335 transcripts encoding lysosomal enzymes and immune-related gene products (Fig. 4A).  
336 At 48 h pf, PROP induced down-regulation of transcripts for shell biogenesis and for  
337 putative neuroendocrine signaling at 0.01 µg/L, and up-regulation of transcripts for  
338 lysosomal and antioxidant responses at 1 µg/L (Fig. 4A). DISTLM showed that expression  
339 profiles strongly depend on the time post-fertilization (89% total variation), with PROP  
340 treatment accounting for about 8% total variation, which mostly explained the observed  
341 transcriptional changes at 48 h pf (Fig. 4B). DISTLM analysis by functional groups showed  
342 that expression patterns of transcripts involved in antioxidant response, immune  
343 responses, shell biogenesis, and, to a lesser extent, lysosomal responses, explained the  
344 observed effects of PROP at 24 h pf (Fig. 4C). The effects of PROP at 48 h pf were mostly  
345 explained by expression patterns of transcripts involved in antioxidant and immune  
346 responses, while putative neuroendocrine signaling and shell biogenesis were of minor  
347 relevance (though statistically significant) (Fig. 4C).

#### 348 4 **Discussion**

349 The 96-microwell embryotoxicity assay has proven as a sensitive methodology for the  
350 high throughput screening of emerging contaminants, including anti-inflammatory, blood  
351 lipid lowering and antidepressant drugs, on embryo development of *M. galloprovincialis*  
352 (Balbi et al., 2018, 2017a, 2016; Estévez-Calvar et al., 2017; Fabbri et al., 2014). The  
353 present study showed that the antiepileptic drug CBZ and the β-blocker PROP, two  
354 pharmaceuticals widely used in human and veterinary therapies and commonly detected in  
355 marine coastal waters, altered embryo phenotypes, with significant effects observed from  
356 environmental concentrations (in the ng/L range). Although comparable EC50 values were

357 obtained in embryotoxicity tests ( $\approx 1 \mu\text{g/L}$ ), CBZ mainly induced shell malformations in D-  
358 veligers at 48 h pf at all the concentrations tested, whereas PROP resulted in different  
359 percentages of malformed/immature, and progressive developmental arrest/mortality at  
360 concentrations much higher than reported environmental levels.

361 In order to investigate possible impact of CBZ and PROP on gene expression, the  
362 effects of both compounds on the transcriptional pattern of 17 genes were evaluated in  
363 embryos at both 24 h and 48 h pf. These include both gene sequences corresponding to  
364 known biological functions in adult mussels and embryos (Balbi et al., 2017a, 2016), as  
365 well as genes whose transcription has been shown to be modulated by exposure to other  
366 emerging contaminants in mussel early larval stages (Balbi et al. 2016, 2017a, 2018;  
367 Capolupo et al., 2018). Moreover, expression of genes related to lysosomal function was  
368 evaluated in this study.

369 Transcriptional effects were investigated at two selected concentrations:  $0.01 \mu\text{g/L}$ , a  
370 value that falls within the range detected in coastal and estuarine waters (see Table S1)  
371 and  $1 \mu\text{g/L}$  (a value around the observed EC50 for both compounds in embryotoxicity  
372 assays). On the whole, the results show that CBZ and PROP induced significant changes  
373 in gene expression at both concentrations and developmental stages.

374 CBZ and PROP affected transcription of shell-specific transcripts at both post-  
375 fertilization times. Previous works indicated that these transcripts are involved in the  
376 homeostasis of carbonate chemistry at the site of calcification (*CA*, *EP*) and in organic  
377 matrix synthesis (*CS*), which control calcification rates and morphology of the shell (Chan  
378 et al., 2018; Kocot et al., 2016; Ramesh et al., 2017). Accordingly, they are strongly up-  
379 regulated in the early stages of mussels along with shell formation (Balbi et al., 2017a,  
380 2016). The down-regulation of both *EP* and *CS* induced by both compounds at 48 h pf  
381 suggests that CBZ and PROP may interfere with regulation of shell biogenesis.



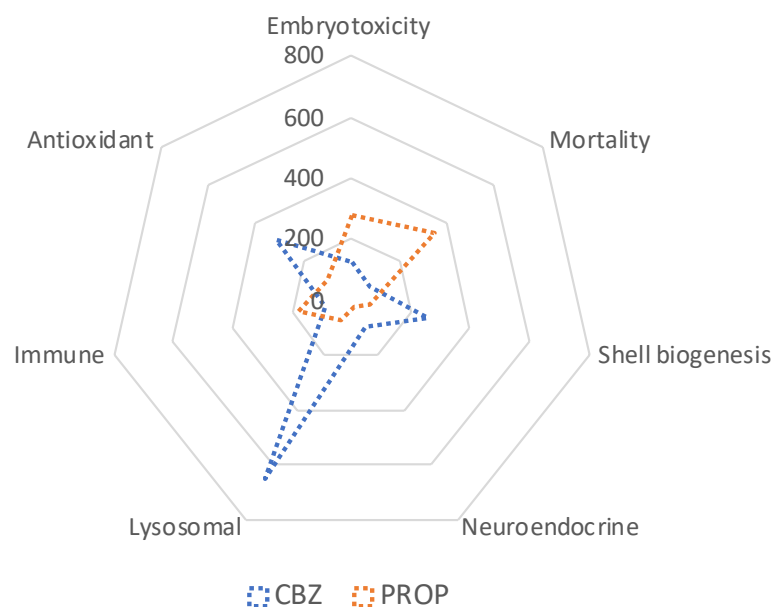
382 Modulation of transcripts encoding for lysosomal enzymes seems a common effect of CBZ  
383 and PROP in developing embryos. In embryos grown in physiological conditions,  
384 expression of lysosomal enzymes *HEX* and *CTSL*, and, to a lesser extent, of *GUSB*  
385 increased dramatically from 24 h to 48 h pf. This is not surprising, since the acquisition of a  
386 suitably functional lysosomal system is related to development of the digestive system,  
387 where intracellular digestion in hepatopancreatic cells represents the main route of food  
388 processing in bivalves (Balbi et al., 2017b), and also endows the animals with an active  
389 detoxification system of waterborne chemicals (Balseiro et al., 2013). Lysosomes may  
390 accumulate pharmaceuticals (in particular lipophilic compounds with amino groups) by pH  
391 partitioning (Kazmi et al., 2013), leading to altered lysosomal functions and drug-drug  
392 interactions (Franzellitti et al., 2015, 2013). The capacity of CBZ and PROP to affect  
393 lysosomes in marine invertebrates was demonstrated by *in vivo* studies with adult  
394 mussels, clams, and crabs, in which the pharmaceuticals decreased lysosomal membrane  
395 stability in haemocytes (Aguirre-Martinez et al., 2015; Aguirre-Martínez et al., 2013b,  
396 2013a; Franzellitti et al., 2011; Martin-Diaz et al., 2009), or increased neutral lipid contents  
397 in digestive glands, and interactively affected PROP tissue accumulation (Franzellitti et al.,  
398 2015). Both CBZ and PROP affected transcription of genes encoding for different  
399 lysosomal enzymes, in particular at 48 hpf. Although the effects were not apparently  
400 concentration-dependent, the results show that the lysosomal function can be affected by  
401 pharmaceuticals also in developing embryos.

402 CBZ and PROP affected transcription of genes involved in immune and antioxidant  
403 responses, causing up- or down-regulations depending on concentration and the stage  
404 investigated. In particular, at 24 hpf, immune-related genes were up-regulated in response  
405 to low concentrations of CBZ. Both immunomodulatory and pro-oxidant effects of

406 pharmaceuticals have been widely reported in aquatic species (Fabbri and Franzellitti,  
407 2016).

408 Finally, CBZ and PROP also affected transcription of serotonin and *Mytilus* estrogen  
409 receptors (*5-HT1*, *MeER1* and *MeER2*). A putative role for these genes in neuroendocrine  
410 signaling during development was suggested by their strong upregulation from fertilization  
411 to 48 h pf under basal conditions (Balbi et al., 2016). Moreover, their transcription was  
412 affected by exposure to different types of contaminants (i.e. bisphenol A, 17 $\beta$ -estradiol,  
413 and styrene) with up and down regulations depending on the chemical and developmental  
414 stage (Balbi et al., 2016; Rasika Wathsala et al., 2018). In particular, the observed 5-HT1  
415 downregulation induced by PROP at the different concentrations and times pf tested,  
416 indicate an overall impact of this compound on the serotonergic system.

417 In an attempt to compare the overall impacts of CBZ and PROP on mussel embryos, a  
418 radar plot that summarizes all data on phenotypical and molecular changes is reported in  
419 Fig. 5.



428

**Fig 5. Radar plot summarizing the embryotoxic and molecular effects of CBZ and PROP in *M. galloprovincialis* early larval stages.** For each biological endpoint, CBZ or PROP concentration-related variation is expressed by the Area Under the Curve (AUC) according to Franzellitti et al. (2018). Details for AUC calculation are reported in Table S5, Supplemental material. **Colored figure is intended only for the online and PDF version.**

429 Although CBZ showed more limited phenotypic outcomes at increasing  
430 concentrations, signatures of molecular effects (antioxidant processes, shell biogenesis,  
431 lysosomal function) which may forecast upcoming changes at the physiological level are  
432 highlighted. For example, the observed changes in lysosomal related gene products in  
433 CBZ-exposed embryos may affect not only the development of the digestive system, but  
434 also degradative pathways supported by lysosomal enzymes that are crucial in embryo  
435 remodeling across development (Dyrynda et al., 1995).

436 Differently, PROP seems to have a greater impact on development in terms of  
437 embryotoxicity and mortality at higher concentrations. Developmental impairments due to  
438 PROP were previously reported in oysters; these effects were partly ascribed to the  
439 therapeutic action of PROP as an adrenergic antagonist, which may alter the physiological  
440 function of catecholamines in mussel embryo development (Yang et al., 2014). Indeed,  
441 PROP blocked epinephrine-induced metamorphosis in larvae of *Crassostrea gigas* (Coon  
442 and Bonar, 1987), increased mortality rates in zebrafish embryos, and induced  
443 developmental arrest in sea urchin embryos (Ribeiro et al., 2015). In embryos of the sea  
444 urchin *Arbacia lixula*, reduced cholinergic and serotonergic signaling was related to  
445 impaired skeletogenesis and increased morphological abnormalities (Cappello et al.,  
446 2017). In this light, additional investigations are needed to understand the effects of PROP  
447 on adrenergic signaling in bivalve development.

## 448 **5 Conclusions**

449 Overall, the present study supports the use of the 96 microwell embryotoxicity assay  
450 to identify those pharmaceuticals that can provoke major types of morphological  
451 alterations at environmental concentrations.

452 With regards to data obtained on gene expression, the results indicate that the main  
453 effects on overall transcription (i.e. maximum distance between control and treated

454 samples in the DISTLM analysis) were detected at 24 h pf for CBZ and at 48 h pf for  
455 PROP. In particular, as shown by the heatmap analysis, the impact of CBZ was evident at  
456 concentrations as low as 0.01 µg/L.

457 The results as a whole do not reveal a clear pattern in response to different  
458 concentrations of CBZ and PROP. In this light, present data do not provide evidence for a  
459 major or distinct mechanism of action for either compound, but they may reflect an  
460 adaptation response to exposure to these pharmaceuticals.

461 However, these are the first data demonstrating that both CBZ and PROP, at  
462 concentrations encompassing environmental levels, can interfere with multiple processes  
463 in mussel early development. Together with recent studies (Balbi et al., 2018, 2016), these  
464 observations raise new concerns on the occurrence of pharmaceuticals in coastal  
465 environments and their potential impacts on marine fauna. This information can be used to  
466 address further investigations on the possible molecular targets and mechanisms of action  
467 for either compound, and eventually contribute to establish regulatory priorities to limit the  
468 environmental occurrence of these emerging contaminants.

469

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