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Toxic Chemical Governance Failure in the United States: Key Lessons and Paths Forward

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Title:

Toxic Chemical Governance Failure in the USA: Key Lessons and Paths Forward

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Abstract:

Over 40 years of regulations in the United States have failed to protect human and environmental health. We contend that these failures result from the flawed governance over the continued production, use, and disposal of toxic chemicals. To address this failure, we need to identify the broader social, political, and technological processes producing, knowing, and regulating toxic chemicals, collectively referred to as toxic chemical governance. To do so, we create a conceptual framework covering five key domains of governance: knowledge production, policy design, monitoring and enforcement, evaluation, and adjudication. Within each domain, social actors of varying power negotiate what constitutes acceptable risk, creating longer term path dependencies in how they are addressed (or not). Using existing literature and five case studies, we discuss four paths for improving governance; evolving paradigms of harm, addressing bias in the knowledge base, making governance more equitable, and overcoming path dependency.

Keywords: toxic chemicals, governance, innovation, path dependency, coproduction

1 **Introduction:**

2 Despite decades of legislation (Wagner 2007) and public interest litigation (Lind 2015),
3 contemporary society remains saturated with environmental pollution risks of its own production
4 (Beck 2008). Globally, these failures of environmental protection cause millions of premature
5 deaths per annum and cost society billions of dollars in economic damages (Landrigan et al. 2017).
6 These pervasive and increasing environmental threats often remain unknown until publicized by
7 private citizens, scientists, NGOs, or the media. Warnings issued with increasing frequency by the
8 scientific community (e.g., Ripple et al. 2017), are all too often met with policy gridlock and a lack
9 of substantive government action. Despite the existence of cleaner technologies whose economic
10 and social benefits exceed transition costs, globally, environmental pollution has become the
11 leading cause of preventable death (Landrigan et al. 2017).

12 Globally, significant gains have been made in reducing primary emissions of some highly
13 regulated chemicals (SC 2017), although debate continues as to the overall impacts of shifting
14 global geographies of the production of toxic risks (Rasli et al. 2018). The US serves as an
15 excellent case study on the multi-faceted nature of governing toxic chemical risks. It has lagged
16 behind the EU (European Union) in adopting the precautionary principle, especially with regards
17 to importing consumer and industrial products (Becker 2010), and in dealing with emerging
18 contaminants (Bao et al. 2015). And as elsewhere, partial solutions have led to unintended
19 outcomes, such as increases in ozone exposure concomitant with declines in particulate pollution
20 due to widespread adoption of catalytic converter technology (HEI 2019).

21 As an interdisciplinary group of scholars crossing the domains of environmental science,
22 public administration, and political science, we attribute failures within the United States to a more
23 general failure of environmental governance (Mol 2016). A focus on governance highlights how

1
2
3 1 different domains of social decision-making define and manage risks and responsibilities
4
5 2 associated with the production and distribution of toxic substances. Governance also centers the
6
7 3 long running concerns of professional vs. lay person knowledge (Brown 1992, Brulle and Pellow
8
9 4 2006), contestations over ‘facts’ in the ‘post truth’ era, and the degree to which administrative
10
11 5 power can shift regimes of environmental governance (Revesz 2019).
12
13

14
15 6 Existing work documents how failures of environmental governance results in
16
17 7 environmental injustice through the inequitable distribution of exposure to toxic chemicals based
18
19 8 upon racial and socio-economic identities (Landrigan et al. 2017). This unevenness has resulted
20
21 9 from racist and opportunistic practices of uneven permitting and enforcement (Morello-Frosch
22
23 10 and Shenassa 2006) and contributes to the framing of governance failures as ‘somebody else’s
24
25 11 problem’ (Pastor and Morello-Frosch 2018). Simultaneously, toxic chemical risks are ubiquitous
26
27 12 and systemic in nature, affecting humans across the globe regardless of their socio-economic class
28
29 13 (Schwarzenbach et al. 2010). Existing support for high environmental quality across the political
30
31 14 spectrum (Feinberg and Willer 2013), combined with rising rates of developmental and chronic
32
33 15 diseases (Landrigan et al. 2017), indicates that there is an urgent need to frame both risks and
34
35 16 policy proposals in a way that mobilizes those of diverse political orientations.
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40 17 The current political climate in the USA indicates significant resentment against the
41
42 18 political establishment, typified by a resurgent anti-administrative state agenda reminiscent of the
43
44 19 1980s (Hejny 2018) and significant negative consequences for public and environmental health
45
46 20 (Cutler and Dominici 2018). On the upside, the current administrative swing has exposed the long-
47
48 21 standing pattern of elite interests disproportionately writing, lobbying, and adjudicating
49
50 22 environmental laws in their narrowly defined self-interest, and has increased mobilization of
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52 23 NGOs, community based organizations, and science-based advocacy organizations (Mol 2016).
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3 1 This political landscape highlights a need for scientists to engage directly with increased public
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5 2 scrutiny (Latour 2004) by calling for democratic governance to employ best available knowledge
6
7 3 for protecting the quality of our environment and public health.
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10 4 With this goal in mind, we provide a conceptual overview of the governance structure of
11
12 5 toxic risk management in the United States. We use our conceptual framework to analyze several
13
14 6 high-profile case studies, and discuss a proposed set of principles, ongoing initiatives, and
15
16 7 challenges of improving toxic chemical governance in the USA.
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19 8

20 21 9 **Toxic Substances Policy in the USA**

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23
24 10 Literature in the United States has documented numerous instances of failure across diverse
25
26 11 classes of pollutants, natural systems, and regulatory contexts (e.g., Davies and Mazurek 1998,
27
28 12 Paavola 2006, Fletcher 2009). Common causes of attributed failure include a failure to regulate
29
30 13 classes of toxic chemicals (Mesnage et al. 2015), standards inadequate to achieve protection
31
32 14 (Vogel and Roberts 2011, Boone et al. 2014), and non-enforcement of existing regulations (Farber
33
34 15 1999). We define failure as unacceptable levels of human and environmental exposure to toxic
35
36 16 chemicals during their production, use, transport, and fate in the environment.
37
38

39
40 17 While existing regulations and policies written by legislatures and enacted by executive
41
42 18 and administrative branches of government (e.g., federal, state, tribal, and local agencies)
43
44 19 ostensibly act in the public interest, other social actors actively shape their design and language
45
46 20 (e.g., lobbying from industry and citizen groups; Davies and Mazurek 1998, Cash et al. 2006) to
47
48 21 constrain their effectiveness. Additionally, manufacturers, installers, and users of potentially toxic
49
50 22 substances routinely evade effective regulation through legal and illegal means (Lynch and
51
52 23 Stretesky 2014). In response to these recognized drivers of failure, remediation efforts generally
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3 1 prioritize stricter regulation based upon the perceived risks of the substance in question. Tactics
4
5 2 include limiting harmful exposure and environmental releases via command and control
6
7 3 regulation, mitigating ongoing exposures with funds generated from regulation, and, in some
8
9 4 cases, providing incentives for eliminating sources of risk by shifting to alternative technologies
10
11 5 (Wilson and Schwarzman 2009). However, the technological capabilities of manufacturing often
12
13 6 evolve faster than their regulatory apparatuses, and industries themselves have built up a
14
15 7 technological, intellectual, and regulatory ecosystem that has effectively excluded many ‘greener’
16
17 8 technologies (Woodhouse 2006).

18
19
20
21 9 In the United States, the current policy framework around toxic substances remains highly
22
23 10 fragmented among jurisdictions of federal agencies such as the Environmental Protection Agency
24
25 11 (EPA), U.S. Department of Agriculture (USDA), Food and Drug Administration (FDA), and
26
27 12 Department of Health and Human Services (DHHS) (see Table 1). Some states have additional
28
29 13 regulations, such as California’s Proposition 65, requiring the state to publish and annually update
30
31 14 a list of known chemical carcinogens or reproductive toxicants (Nelson 2013). Mirroring
32
33 15 jurisdictional fragmentation resulting from sector-specific regulations, variation exists for different
34
35 16 media (e.g., soil, air, water; Caliman and Gavrilesco 2009, Rudel and Perovich 2009).

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40 17 In addition to poor policy design and fragmentation, the current policy framework leaves
41
42 18 many chemicals un- or under-regulated. The primary federal toxic chemical regulation, the Toxic
43
44 19 Substance Control Act (TSCA; implemented in 1976), has grandfathered in nearly 62,000
45
46 20 previously unregulated chemicals without evaluation of risk (Vogel and Roberts 2001), a number
47
48 21 not including the manufacturing by-products of those chemicals or their environmental derivatives.
49
50 22 A hard-fought 2016 amendment to TSCA established a schedule for evaluating the estimated
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52 23 85,000 existing chemicals in the marketplace, shifted toxicological analyses towards a risk-based
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3 1 framework, limited the ability of companies to claim commercial confidentiality, and has
4
5 2 eliminated the consideration of cost in risk assessment (Frank R. Lautenberg Chemical Safety Act
6
7 3 2016). With a risk-based framework, the burden of proof for evaluating potential harms to humans
8
9 4 and the environment is placed on the regulatory agency, who will only regulate a chemical if it is
10
11 5 shown to pose a risk to human and environmental health in a highly specified exposure pathway.
12
13 6 TSCA remains in litigation over fundamental procedural issues, including the process of
14
15 7 prioritizing different substances for evaluation, definitions of 'unreasonable' risk, and whether the
16
17 8 EPA should consider the feasibility of replacement substances in prioritization (Bergeson and
18
19 9 Graham 2017, CW 2019). Despite the 2016 TSCA requirements for EPA to evaluate all new
20
21 10 chemicals before market release, EPA remains underfunded and understaffed for timely
22
23 11 evaluation. Exacerbating the situation, TSCA does not require companies to provide toxicological
24
25 12 data, and the annual evaluations of 20 high-risk and 20 low-risk chemicals can not keep pace with
26
27 13 new chemical production (Botos et al. 2018).

32
33 14 In addition to TSCA, substances that pose threats to human and environmental health are
34
35 15 regulated by a number of other regulatory instruments including the Clean Air Act, Clean Water
36
37 16 Act, Safe Drinking Water Act, Resource Conservation and Recovery Act, and various workplace
38
39 17 regulations under Occupational Safety and Health Administration – all of which rely on similar
40
41 18 processes of analyzing risk to determine the extent to which they should be regulated (Steward
42
43 19 1995). Additionally, chemicals intended for human consumption as food stuffs, pharmaceuticals,
44
45 20 tobacco products or derivatives, and personal care products undergo their own regulatory
46
47 21 procedures through the Food and Drug Administration. Further fragmenting the regulatory
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49 22 environment, the Federal Insecticide, Fungicide, and Rodenticide Act (FIFRA), regulates the sale
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51 23 of agricultural chemicals not registered with the EPA. Though this process requires stringent
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3 1 manufacturer testing, labeling, and periodic recertification (every 15 years) of pesticides, it critical
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5 2 economic importance can outweigh human and environmental risks, especially if those substances
6
7 3 have become widespread, or engineered into crop production systems. Perhaps because of this it
8
9 4 is rare for a pesticide to be denied re-registration unless there is overwhelming evidence of human
10
11 5 and ecological harm. Both FIFRA and TSCA suffer in effectiveness due to their definitions of risk,
12
13 6 and the ease of industry influence on their decision-making processes.
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17 7 In contrast, the European Union's Registration, Evaluation, Authorization and Restriction
18
19 8 of Chemicals (REACH) legislation, utilizes precautionary principles, putting the burden of proof
20
21 9 of safety on the industries that produce them (Silbergeld et al. 2015). Under the precautionary
22
23 10 principle, new substances and their derivatives are assumed to pose risk until proven otherwise,
24
25 11 and the responsibility for proving the absence of risk is placed on both the producer and the
26
27 12 regulator. REACH also requires producers to provide the European Chemicals Agency (ECHA)
28
29 13 with toxicological information, and uses a spectrum of safety standards matched with appropriate
30
31 14 use restrictions and mandatory labeling.
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35 15 All of the above policies, struggle with inadequate resources for chemical assessment and
36
37 16 are vulnerable to lobbying and uneven adjudication. A primary challenge with REACH, for
38
39 17 example, is that EU member states are responsible for implementing the chemical evaluation
40
41 18 process, leading to inconsistencies in implementation. Despite the improvements made to TSCA,
42
43 19 it still falls short of REACH's founding precautionary principles. This comparison indicates the
44
45 20 importance of strong guiding principles in effective toxic chemicals governance. While the
46
47 21 incorporation of precautionary principles will be an important step toward better regulation of
48
49 22 toxic chemicals in the U.S., it alone is inadequate. Effective protection from toxic chemical risks
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51 23 will require changes in governance practices and an evolution of the technologies and practices
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3 1 that generate risk to align with public and environmental health goals.
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3 **A Framework for Understanding Toxic Chemical Governance**

4 Governance includes the social practice of designating rules, standards, and norms
5 according to which actors and institutions negotiate and make decisions (Rogers and Hall 2003),
6 including what knowledge is considered valid and useful (Wynne 2003). A focus on governance
7 identifies how the present conditions of our society, environment, and technological
8 infrastructures are interdependent with the forms of expertise and political authority deemed
9 necessary to manage harms to humans and the environment (Scott 1998, Jasanoff 2004, Latour
10 2004). In particular, governance highlights the process of classifying potentially hazardous
11 substances as risks; the consideration of different forms of knowledge or expertise in that
12 process, and the path dependency – or inertia – resulting from prior decisions. We describe these
13 concepts below and use them later to evaluate five high profile cases of toxics regulation success
14 or failure in the United States.

15 The construction of risk around toxic chemicals can be defined as a social process of
16 emphasizing some dangers over others (Douglas and Wildavsky 1983). Formal risk analysis
17 involves calculating the probability of a specified level of chemical exposure multiplied by the
18 probable consequences of that exposure (Bocking 2004). However, in practice, such analysis
19 relies on a set of assumptions about social behavior alongside physiological and toxicological
20 data, and often disregards risks experienced by affected communities (Bocking 2004, Beck
21 2008). In this sense, standard risk analysis treats risks to public health and the environment as
22 “end of pipe” problems and unplanned releases as public relations problems. Such thinking
23 ignores that the generation of risk results from choices about how chemicals can be produced.

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2
3 1 These choices produce systems which have global consequences, normalize the production of
4
5 2 toxic byproducts, and have significant sunken costs in facilities and the development of
6
7 3 economic sectors dependent upon those kinds of inputs (Beck 2008).
8
9

10 4 While different social actors perceive and calculate risks in different ways, risk
11
12 5 management is generally seen as an activity worthy of professional expertise. Expertise in this
13
14 6 sense refers to the social practices of designating individuals, institutions, technologies, and
15
16 7 methods as sources of authoritative knowledge (Scott 1998, Wynne 2003, Bocking 2004,
17
18 8 Jasanoff 2004). Expertise often has disciplinary boundaries, which prevent synthesis across and
19
20 9 within disciplines (Cartwright 1999). As social actors vie for legitimacy within networked
21
22 10 political, financial, environmental, social, and technical systems (Grabowski et al. 2017)
23
24 11 institutions take on more stable forms, routing social decision-making processes into established
25
26 12 mechanisms and fora exhibiting different forms of path dependency.
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31 13 Path dependency refers to the way in which future possibilities are seen as constrained by
32
33 14 present conditions and largely results from decisions about financial, institutional, intellectual,
34
35 15 and bureaucratic investments in social ways of doing, infrastructures, and technology (Jasanoff
36
37 16 2004, Woodhouse 2006, Beck 2008). Path dependencies may lead to the generation of systemic
38
39 17 bias in what type of knowledge is produced and considered relevant, which is often contested by
40
41 18 popular movements (Hess 2015). Disrupting path dependency generally requires major events, a
42
43 19 form of punctuated equilibrium (Pierson 2000). Systemic path-dependencies result when agents
44
45 20 within institutions prevent change despite widely recognized problems (Sydow et al. 2009). For
46
47 21 example, toxic chemical risks have often been framed as by-products or externalities, or
48
49 22 attributes of chemicals to be managed, when in fact they are embedded within 'normal'
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51 23 operations (Beck 2008, Perrow 1984).
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1 Drawing upon these concepts of risk, expertise, and path dependency, we present a
2 conceptual framework of the current toxic chemical governance system in the USA. We break
3 down the overall governance system into five interdependent domains in which toxic chemical
4 regulations are interpreted, implemented, and evaluated: knowledge production, policy design,
5 monitoring and enforcement, evaluation, and adjudication (Figure 1). Each domain operates
6 simultaneously in time and space, although problems can flow from one to another (e.g., failures
7 of enforcement often result in adjudication).

8 Knowledge Production

9 While different forms of expertise and knowledge are embedded in all domains, the
10 ‘knowledge base’ refers to the overall organization of information pertaining to toxic chemicals.
11 This includes “facts” and “information,” and the accepted methods for producing them, which
12 invokes the ways institutions, values, norms, and discourses within a social system decide what
13 type of knowledge is legitimate or useful (Jasanoff 2004, Stehr 2015). Many stakeholders are
14 involved in toxic chemical knowledge creation, including affected communities, the scientific
15 community, the media, and industry representatives. Each stakeholder group constructs their
16 knowledge differently, leading to different claims about toxic chemicals. These varied claims
17 and perspectives on what constitutes legitimate knowledge, and how it is and should be
18 produced, lead stakeholders to identify and categorize threats to health and the environment in
19 radically different and often incompatible ways (Wynne 2016).

20 Policy Design

21 Policy design includes processes for describing present conditions, framing goals,
22 creating incentives and/or regulations to achieve those goals, and assigning rights and

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2
3 1 responsibilities to different social actors within an overarching policy architecture. This domain
4
5 2 heavily influences monitoring, enforcement, and evaluation activities, and sets the stage for
6
7 3 adjudication. It is here that the interests and values negotiated within the knowledge base
8
9 4 become codified into legislation via regulations, incentives, and budget allocations. Top-down
10
11 5 policy architecture is more easily implemented but less flexible for local stakeholders, whereas
12
13 6 bottom-up approaches are adaptive and flexible, but can be difficult to create given disagreement
14
15 7 among stakeholders and/or lead to unequal environmental regulations across the country
16
17 8 (Bocking 2004). Canonical descriptions of the policy process divide participants into decision-
18
19 9 makers, generally referring to elected officials, and stakeholders including: affected
20
21 10 communities, industries, and ‘special interest’ groups. It has been observed that local affected
22
23 11 communities engaged in the policy process often demand a precautionary approach to protect
24
25 12 their local human and environmental safety (e.g., Bullard and Johnson 2009), while industry
26
27 13 interests push for limiting regulation and including policy language that allows them to continue
28
29 14 current business operations (e.g., Boone et al. 2014).

36 Monitoring and Enforcement

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38
39 16 Monitoring and enforcement refers to the mechanisms of observing regulated activities
40
41 17 and the ability to coerce compliance with standards and operating procedures as written.
42
43 18 Enforcement can take place via three primary approaches. (1) The formal regulatory arena: local,
44
45 19 state, and federal executive and regulatory agencies issue fines for limit exceedances and issue
46
47 20 release permits, among other codified approaches to compliance. This requires sufficient
48
49 21 resources for detecting and correcting violations. (2) Self-regulated monitoring and enforcement:
50
51 22 in the absence of close regulatory oversight, private-contract auditing agencies oversee industry
52
53 23 groups to ensure compliance, often via certification programs. (3) Complaints by affected
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1 communities who identify misconduct, draw media attention, and place political pressure on
2 industry groups to comply with regulations. This often happens when there are limitations in
3 agency resources. Proper enforcement requires adequate policy design, including initial political
4 will, coherence in writing legislation, and consistent, long-term political and financial support for
5 monitoring and enforcement efforts (Wagner 2007).

6 Programmatic and Policy Evaluation

7 Programmatic and policy evaluation refers to evaluation of policies designed to manage
8 chemical exposure, production, and transportation, and the creation of alternative technologies
9 and practices. This domain is tightly linked to policy design, whereby program and policy
10 evaluations should inform future policy designs. Major actors involved in this domain include
11 federal, state, and local regulatory government agencies, affected communities, industries, the
12 scientific community (including non-governmental agencies and nonprofits), and the media.
13 Judicial agencies that perform policy evaluations in response to publicity of toxic chemical risk
14 to human or environmental health, and/or to initiatives within industry to change their practices
15 or use of certain chemicals may also be involved. This is the domain in which stakeholder claims
16 about the impacts of policies are evaluated, then either utilized to evolve policy or disregarded.

17 Adjudication

18 Adjudication is the legal process by which disputes are settled, policies are interpreted
19 (e.g., claims of harm and liability), and enforcement activities are contested (e.g., ongoing TSCA
20 litigation pertaining to procedural rules for chemical risk evaluation). A key part of adjudication
21 pertains to the formal determination of compliance, liability, harm, and responsibility to parties
22 involved in litigation.

1
2
3 1 More broadly, adjudication is the process by which knowledge claims, policy efficiency,
4
5 2 and the distribution of benefits/burdens of a particular substance are vetted by the judicial branch
6
7 3 of government. Affected and/or scientific communities, non-governmental interest groups,
8
9 4 government agencies, and industries often initiate adjudication (Hoffman 1999) as a means of
10
11 5 changing activities within the other governance domains. These changes may include policy
12
13 6 design modifications, enforcement of compliance, increased monitoring and enforcement, and
14
15 7 promoting or contesting evaluation. Common forms of adjudication include petitions to state and
16
17 8 federal agencies, settlements with regulatory and private sector entities, or lawsuits. While the
18
19 9 courts can settle issues of human and environmental failures, this approach is, by definition,
20
21 10 reactionary and can only interpret legislation to nullify/clarify obligations, or set appropriate
22
23 11 enforcement actions, through slow, costly, and often adversarial means (Silbergeld et al. 2015).
24
25 12 Adjudication can pre-emptively impact policy as legislatures shy away from creating un-
26
27 13 enforceable policies.
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33 14

35 15 **Application of Governance System Conceptual Framework**

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37 16 We expand our conceptual framework of five governance domains (i.e., knowledge
38
39 17 production, policy design, monitoring and enforcement, evaluation, and adjudication) to
40
41 18 determine how patterns of flawed governance lead to unsafe exposure of select chemicals. We
42
43 19 create a qualitative evaluative framework for governance issues related to risk definition,
44
45 20 knowledge production, and path dependency across the five domains (Figure 2). Based on
46
47 21 aspects of the governance literature discussed above, we consider whether risk, expertise, and
48
49 22 path dependency are “succeeding,” “failing,” or “partially succeeding”, corresponding with a
50
51 23 numerical ranking (see Figure 3). Risk is qualitatively evaluated as succeeding if there is
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1 plurality and consensus of how the risk is framed and it is considered failing if risk is understood
2 only from one perspective and/or is highly contested. Expertise is qualitatively evaluated as
3 succeeding if there are multiple forms of knowledge and participants in addressing the risk,
4 including those most affected by the decision; it is considered failing when only one form of
5 expertise or knowledge is used and the perspectives of affected parties are disregarded. Path
6 dependency is considered succeeding if the system is evolving to address emerging risks and
7 challenges and is considered failing if it is regressing or failing to evolve despite an
8 acknowledged need to do so.

9 We then apply this evaluative framework (Figure 2) to five high-profile case studies. We
10 chose a set of toxic substances case studies based upon their representativeness within
11 infrastructure systems (lead and SO_x), manufacturing (heavy metals in light industry),
12 agriculture (glyphosate), and consumer products (bisphenol-a (BPA)). For each case, we
13 assembled literature reviewing the evidence base for each case, collected popular media accounts
14 describing the policy responses, and examined relevant legislation and enabling policies of their
15 regulation (Supplemental Materials 1). Each co-author described the conditions of a case by each
16 domain, and then used the subjective scoring system to rank the robustness of risk framing,
17 representativeness of expertise, and the degree of path-dependency per the evaluative framework
18 in Figure 3. For each case, we averaged the group's scoring for risk, expertise, and path-
19 dependency within each governance domain to compare the perceived level of success or failure
20 between cases, and then discussed the group's findings to achieve consensus on a final ranking.
21 Applying this evaluative framework draws out where failures in toxics governance are rooted in
22 each case, and where there are similarities and differences across cases.

23

1 **Key findings**

2 Most cases ranked between “failing” (score = 1) and “partially succeeding” (score = 2) out of a
3 total possible score of “success” (score = 3) (Figure 2). Overall, we found that although most
4 cases had robust risk knowledge that included diverse perspectives for framing risk, ongoing
5 issues with path dependency, and in some cases policy regression, were common throughout the
6 cases evaluated.

7 Lead in school drinking water

8 Score: 1.4/3

9
10 Schools serving children across socioeconomic strata nationwide have unsafe lead levels
11 in drinking water fountains (Wines et al. 2016), because the Safe Drinking Water Act regulates
12 water lead levels at water treatment facilities, but not at the tap. Updated ‘lead-free’ plumbing
13 rules maintain allowable lead content, and legacy plumbing and water infrastructure management
14 can cause significant lead leaching. Well-publicized cases include schools in Washington, DC in
15 2000 and 2004, Seattle, WA in 2004, Flint, MI in 2014, and Newark, NJ, New York, NY and
16 Portland, OR in 2015 and 2016. The persistence of this issue is caused by failures within the
17 enforcement, monitoring and evaluation, and adjudication domains, along with path dependences
18 within all domains (Figure 4a).

19 Heavy metals in light industry

20 Score: 1.8/3

21
22 Bullseye Glass in Portland, Oregon creates art and architectural glass products. Because
23 of regulatory exemptions for small-scale industry, they lacked scrubbers and were releasing
24 heavy metals, including known carcinogens like cadmium and arsenic (Donovan et al. 2016).

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2
3 1 Portland residents filed eight complaints across multiple decades to the state Department of
4
5 2 Environmental Quality, but no action was taken until the media was notified that the US Forest
6
7 3 Service found levels of cadmium almost 50 times above Oregon's benchmark of 0.6 ng/m³
8
9 4 during a moss air monitoring project (Donavan et al. 2016). A class action lawsuit against
10
11 5 Bullseye Glass was subsequently filed, and a cease and desist order was issued for any
12
13 6 uncontrolled furnaces. Regulatory gaps are now addressed by the Cleaner Air Oregon initiative,
14
15 7 but still exist at the federal level. Based on our evaluative framework, we determined that risk
16
17 8 evaluation for this case was moderately successful, and that the release of unsafe levels of heavy
18
19 9 metals can mainly be attributed to failures of expertise and path dependency (Figure 4b).
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26 11 Sulphur and nitric oxides

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28 12 Score: 2/3

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31 13 Sulfur oxides (largely SO₂) result from burning sulfur or sulfur-containing materials,
32
33 14 mostly coal, but present in all fossil fuels. Nitric oxides (NO_x), which form during hydrocarbon
34
35 15 combustion under an excess of oxygen, are both harmful to human health (affecting respiratory,
36
37 16 cardiovascular, and neurological systems) and the built and natural environment as the leading
38
39 17 causes of acid rain and deposition (Likens 1974, McCubbin and Delucchi 1999). Risk of unsafe
40
41 18 exposure still exists due to industry influence on policy design and monitoring and enforcement,
42
43 19 and due to failures of expertise and path dependency within the policy evaluation and
44
45 20 adjudication domains (Figure 4c).
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49 21 50 22 Bisphenol-a in consumer products

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53 23 Score: 1.3/3
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3 1 This endocrine disruptor (compound that interferes with proper hormone signaling) is
4
5 2 present in many plastics to mitigate brittleness. BPA became notorious when scientists identified
6
7 3 that it can leach into food and drinks and onto skin, and consumption or absorption of this
8
9 4 compound, especially early in development, may increase cancer risk due to its endocrine-
10
11 5 disrupting properties (e.g. Seachrist et al. 2016). Despite these findings, the use of BPA is still
12
13 6 allowed in most products, although market pressure has resulted in its phase-out, and the FDA
14
15 7 has removed it from the list of allowable additives in baby and children's food and drink
16
17 8 products. While some risk has been mitigated thanks to moderately successful knowledge
18
19 9 production and adjudication, failures are still pervasive around proper monitoring and
20
21 10 enforcement due to privileging of expertise and systemic path dependency and the use of
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23 11 replacement chemicals with uncertain toxicity (Figure 4d).
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31 Glyphosate in agriculture

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33 14 Score: 1.4/3
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35 15 Glyphosate is the active ingredient in Roundup ©, one of the most commonly applied
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37 16 pesticides in the United States, where laws require “reasonable certainty of no harm” as a
38
39 17 prerequisite for pesticide certification. Industries typically determine such risk by assessing
40
41 18 health effects at increasingly higher doses (dose-response); however, scientists have found that
42
43 19 low, environmentally relevant concentrations of glyphosate can mimic and interfere with
44
45 20 hormone signaling (endocrine disruption) and may also be associated with non-Hodgkin’s
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47 21 lymphoma (Mesnage et al. 2015). Despite this growing body of scientific literature on the risks
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49 22 of glyphosate, the threshold for maximum glyphosate residues on food and animal feed - known
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51 23 as the tolerance level - continues to increase, and glyphosate was recertified for use in 2015
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1 (Benbrook 2016). The persisting risk of glyphosate exposure can be attributed to pervasive
2 failures across all governance domains, particularly with respect to policy design, monitoring and
3 enforcement, and evaluation. Industry influence over the governance processes has continued to
4 affect ongoing adjudication processes, although harmed individuals have achieved some post
5 harm compensation (Figure 4e).

7 **Discussion: Paths forward**

8 Our analysis of a diverse set of failures to protect public and ecological health from toxic
9 risks indicates a strong need to improve the overall governance of toxic chemical production and
10 use throughout the United States. Four major patterns emerge from our analysis of governance
11 failures:

- 12 1) Governance allowing the production and release of toxic chemicals with inadequate
13 assurance of safety leads to inevitable harm to human and environmental systems,
- 14 2) Certain forms of knowledge, particularly those that favor industry over public and
15 environmental health, are privileged when assessing the extent and risk of this harm,
- 16 3) Knowledge inequality is exacerbated by unequal formal mechanisms for resolving
17 disputes over the assessment, mitigation, and redressing of harms, and
- 18 4) Path dependency of technological, administrative, and knowledge-producing systems
19 makes effective change difficult and perpetuates harm, despite regulatory action.

20 For each of these interrelated issues, we identify prominent paths forward based on a re-
21 interpretation of the purpose of toxic chemicals governance, provide examples of developing real
22 world initiatives addressing them, and discuss challenges to their continued development and
23 success.

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3 1 Issue 1: Incomplete paradigms of mitigation and risk management: the inevitability of harm from
4
5 2 toxic chemical production
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8 3 It is clear from our case studies of lead, glyphosate, SO_x, heavy metals, and BPA that
9
10 4 many toxic chemical risks are persistent both in their sources and their biological consequences.
11
12 5 These risks often only become known after enough harm has accrued to communities to elicit a
13
14 6 social response (Mesnage et al. 2015, Silbergeld et al. 2015). Patterns of enforcement and
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16 7 adjudication indicate that present regulatory processes generally only mitigate or act
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18 8 retroactively, not preventatively. At the current rate of evaluation under TSCA, new chemicals
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20 9 are being manufactured faster than existing chemicals are being evaluated, especially those
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22 10 produced outside of the USA (Bernhard et al. 2017). Even for regulated chemicals that have
23
24 11 reporting requirements, existing datasets fail to communicate the frequency of chemical
25
26 12 exposures or releases that are occurring, leading to enforcement failures as evidenced by the
27
28 13 pervasive presence of toxic chemicals in global ecosystems and human populations
29
30 14 (Schwarzenbach et al. 2010, Bernhardt et al. 2017). Good governance should therefore
31
32 15 incorporate a paradigm shift around toxic chemicals management from one of mitigating risk to
33
34 16 one of eliminating risk and supporting clean production to improve the long-recognized need for
35
36 17 coordinated global and regional governance (Vogel 1997).
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43 18 *Principle 1. The right to be free from toxic chemical risks*
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46 19 Enshrining the right to be free from harm from toxic chemicals in policy will provide a
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48 20 clear articulation of our overall goals as a society with regards to what rights are sacrosanct and
49
50 21 which can be negotiated (Hayward 2002). Ambitious policy goals of eliminating the production
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52 22 of toxic chemicals and supporting the right of humans and ecosystems to be free from harm
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54 23 caused by toxic chemicals will enable transformation of the complex systems producing toxic
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3 1 chemical risks (Jasanoff 2004, Woodhouse 2006). Given the economic benefits that
4
5 2 industrialized countries have already realized by engaging in compliance-based environmental
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7 3 regulation (Wallace 1995), further benefits could be realized by addressing the interdependent
8
9 4 threats of anthropogenic climate change and global pollution, all while revitalizing US
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11 5 manufacturing and providing millions of jobs in the process (Bain et al. 2016).
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15 6 To overcome the significant political, economic, and technological inertia of addressing
16
17 7 these interdependent threats, we can look to the precedent of using purity as a rhetorical tool for
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19 8 political mobilization and for overcoming industry special interests (Barkan 1985). Existing
20
21 9 research recognizes a high degree of support for protecting environmental purity and human
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23 10 health across the political spectrum, despite ideological differences over the role of government
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25 11 in regulating businesses, requiring sustained public mobilization to enact significant legislative
26
27 12 reform (Feinberg and Willer 2013). While such mobilization can set a legislative agenda for
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29 13 technological and economic evolution, a need remains for generating knowledge to enable
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31 14 systemic transformation (McCormick and Kautto 2013).
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34 35 36 15 Issue 2: Biased and Incomplete Knowledge 37

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39 16 Across our cases, we observed a consistent privileging of certain forms of knowledge in
40
41 17 defining and managing risks, which generally favors biophysical laboratory science over field
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43 18 observation, including epidemiological, anthropological, and social science accounts of
44
45 19 experienced risk and harm. Even after harm becomes known, industry and responsible parties
46
47 20 will consistently challenge accounts of harm while hiding behind the same scientific uncertainty
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49 21 that would cast doubt on their initial risk assessments. This tactic is present at the forefront of
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51 22 litigation over glyphosate and lead in school drinking water. Our case studies mirror larger
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53 23 systemic problems in risk assessment, including affiliation bias in the risk assessment arena
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3 1 (Slovic 2016), targeted attacks on independent researchers (Reeves 2015), and the large volume
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5 2 of industry-sponsored toxicological risk assessments (Hartung 2009).
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8 3 Knowledge production around toxic substances in the USA remains fragmented by the
9
10 4 physiochemical and toxicological properties of regulated materials. As chemical classes affect
11
12 5 different exposure pathways, placement of chemicals within the overall economic system (e.g.,
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14 6 during their production, release in the environment, or use in consumer products) is an important
15
16 7 consideration in proper regulation. However, there is little systematic coordination in the
17
18 8 production of knowledge of contaminant classes based upon their chemical structure and mode
19
20 9 of action, or how these classes are used and released. Current toxic chemical governance uses a
21
22 10 narrow approach to knowledge production instead of, for example, evaluating substances based
23
24 11 on classes with shared chemical structure – such as organochlorines or brominated flame-
25
26 12 retardants – or even based on shared mode of actions – such as level or type of carcinogenicity.
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28 13 A class-based approach may lead to more effective and efficient regulation and protection from
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30 14 chemical risks (e.g., Sanderson et al. 2004), and has been partially adopted by the current TSCA.
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35 15 The failure to include broad expertise in the governance process has cascading effects:
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37 16 policy design does not adequately prevent failures, and often does not provide architecture for
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39 17 effective monitoring and enforcement. As path dependency is rigid, effective policy evaluation is
40
41 18 often nearly impossible. This means that adjudication is necessary to attempt to address
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43 19 grievances, while effective change is made difficult by poor policy design, lack of monitoring
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45 20 and enforcement, and the institutional challenges to quality policy evaluation, including major
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47 21 limitations on building a knowledge base for alternative chemical production.
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52 22 *Principle 2: Support diverse knowledge systems*
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3 1 Evolving the knowledge base entails supporting the generation and synthesis of diverse
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5 2 forms of knowledge for a more robust understanding of the complex nature of toxic chemical
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7 3 risks and the resulting socio-technical transformations needed. At present, advances have been
8
9 4 made in funding independent evaluations of chemical toxicity and in improving both laboratory
10
11 5 and field-based methods for assessing toxic chemical risks. However, some promising
12
13 6 technologies, such as the use of cell cultures and metabolic micro-arrays instead of animal
14
15 7 testing, could dramatically cut the costs of risk assessment but require sustained investment in
16
17 8 order to penetrate a field dominated by animal testing (Hartung 2009). These advances in
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19 9 laboratory science should also be interdependent with field-based, public health, and experiential
20
21 10 knowledge of toxicity. Increases in knowledge generation and synthesis about the impacts of
22
23 11 toxic chemicals also need to inform and integrate research on alternative modes of clean
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25 12 production for substances of similar function, all which could be funded by implementing fees
26
27 13 on the production of certain chemical classes (Thornton 2000). Building such a diverse
28
29 14 knowledge system is not without its challenges, many of which can be overcome by providing an
30
31 15 inclusive, representative, outcome focused, and independently evaluated research process for
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33 16 different classes of toxic chemicals (Reed et al. 2014). However, as our case studies indicate,
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35 17 integrating diverse knowledges requires substantive changes throughout the rest of the
36
37 18 governance system.

38 Issue 3: Uneven and Unequal Governance

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48 20 In cases where laboratory science presents significant evidence of risk of widely used
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50 21 chemicals, such as BPA, glyphosate, and lead in plumbing, unequal policy and enforcement
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52 22 mechanisms privilege the material interests of powerful actors over the health and well-being of
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54 23 communities and ecosystems. A lack of resources for adequate regulatory enforcement and
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3 1 policy and program implementation is symptomatic of the skewed priorities of the existing
4
5 2 governance system. Some of our case studies exhibited partial success in one or more domains.
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7 3 For example, adjudication in glyphosate under the logic of compensation, allows for continued
8
9 4 operations, serving as a “bandage” to mitigate core weaknesses in policy design and
10
11 5 enforcement. This model of governance disproportionately affects vulnerable populations,
12
13 6 including children, elderly, low-income individuals, and future generations in favor of industry
14
15 7 (Elliott et al. 2004, Landrigan et al. 2017).
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19 8 Part of the reason for this uneven and unequal governance is the influence industry has on
20
21 9 shaping the present policy sphere. These types of failures result from targeting public opinion
22
23 10 (Robbins 2007), and the lobbying and influencing of legislators (Hall and Deardorff 2006,
24
25 11 Fredriksson et al. 2003) to the point where legislation drafted by industry associations can
26
27 12 become law (Potter 2011). This legislative capture is often reinforced by regulatory capture,
28
29 13 occurring when an executive agency meant to protect public interest instead protects the industry
30
31 14 it regulates (Shapiro 2012). Arguments for this close relationship between regulators and
32
33 15 industries hinge on the idea that the two entities are supposed to collaborate to provide economic
34
35 16 growth while protecting public values and interests (Lind 2015). By extension, the relatively
36
37 17 limited influence on the policy process exerted by environmental and public health interest
38
39 18 lobbyists, and their shift towards legal expertise, has resulted in a system of ‘regulation by
40
41 19 litigation’ (EPA 2017) by the ‘public interest law complex’ (Lind 2015). Overall, these tensions
42
43 20 highlight that while some adjudication can lead to substantive enforcement actions, without
44
45 21 significant policy change and associated governance evolution, seeking financial redress from
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47 22 toxic industries may perversely promote increased or dirtier production as companies must
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49 23 finance compensatory penalties from their operating budgets.
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3 1 *Principle 3: Inclusive, transparent, and accountable institutions*
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5 2 Overall, effective governance comes from increasing the representativeness and
6
7 3 transparency of democratic processes, and allowing for the direct involvement of affected
8
9 4 communities in policy design and implementation. Such a principle supports two primary
10
11 5 initiatives: building a collaborative governance body and identifying cross-scale institutional
12
13 6 linkages needed to address the complexity of contemporary global industrialization.
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17 7 The creation of a collaborative governance body may help alleviate some of the patterns
18
19 8 we have highlighted. Successfully building a collaborative governance body involves bolstering
20
21 9 participatory science approaches (e.g., citizen science programs) to narrow the science-policy
22
23 10 gap. Specifically, a collaborative governance body would (1) consider and evaluate traditional
24
25 11 ecological knowledge, scientific knowledge, and the experiential knowledge of affected
26
27 12 communities (e.g., Bäckstrand 2003); (2) include diverse stakeholders in knowledge exchange
28
29 13 (Reed et al. 2014); and (3) engage procedural elements, such as independent moderation, to
30
31 14 ensure a balance of power within the group (Purdy 2012). Collaborative governance is also
32
33 15 mutualistic with collaborative knowledge production, and it decreases monitoring costs and
34
35 16 increases industry accountability while empowering communities (Johnson et al. 2014).
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40 17 Examples of such bodies presently exist, although not without their own challenges. As
41
42 18 with toxic chemicals, our oceans are governed by a diversity of laws, regulations, and agencies.
43
44 19 To address this fractured governance, the National Ocean Policy Act (NOPA) was passed in
45
46 20 2010, establishing the National Ocean Council, a collaborative body that includes representatives
47
48 21 of the federal departments and agencies with major jurisdiction over the oceans to share
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50 22 resources and collaborate to implement policy. The NOPA marks the first national effort to
51
52 23 implement a holistic, multi-agency approach to managing our coasts and oceans, although
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3 1 contrasting priorities between executive administrations have limited its effectiveness (Malakoff
4
5 2 2018). The implementation and challenges of the NOPA indicate that governance of complex
6
7 3 social-environmental and technological systems requires operating horizontally across sectors
8
9 4 (industries, media, academic scientists) and vertically across levels (communities, agencies,
10
11 5 legislatures) (Cash et al. 2006).
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15 16 Issue 4: Path Dependency and Inertia

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18 7 The path dependency exhibited in each case results from the costs sunken into certain
19
20 8 means of production (i.e., technologies producing toxic risks), the persistence of many toxic
21
22 9 materials, privileging of knowledge (Brown 1992), and the general absence of self-corrective
23
24 10 behavior by industries, barring significant social influence. Current market logics enabled by
25
26 11 state regulation have proven inadequate for internalizing the costs of production and have
27
28 12 violated the economic principles of functioning markets (Haldane et al. 2017). More troublingly,
29
30 13 industry priorities have continued to shape research and development towards minimizing costs
31
32 14 and maximizing profits as opposed to alternative means of production (Woodhouse 2006).
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37 15 Additionally, many emergent risks are systemic, in that they emerge from complex
38
39 16 interactions between society, the environment and technologies, such as SO_x and NO_x resulting
40
41 17 from automotive pollution. While small technological fixes like improved catalytic converter
42
43 18 technology and conversion to electric vehicles are possible, the aggregate influence of car-
44
45 19 dependent suburban development has outstripped gains from cleaner combustion technology. At
46
47 20 the same time, innovation in some sectors has been shown to reduce risks from long-entrenched
48
49 21 interests, evidenced by the grid purchasing power parity of wind farms over coal, facilitated by
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51 22 direct investments in research and development, and significant policy support for fledgling
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53 23 industries (Jenkins et al. 2010). A proper innovation-oriented approach can facilitate long-term
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3 1 system evolution, as opposed to the reactionary method of bureaucratizing risk that have led to
4
5 2 systemic path dependencies that undermine sufficient toxic chemical governance.
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9 3 *Principle 4: Invest in innovation and real-world deployment*

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11 4 Policies need to identify and incentivize ways of producing substances that meet the
12
13 5 goals and needs of contemporary society without exposing people to toxic chemicals. Existing
14
15 6 command and control, and other ‘end of pipe’ regulations, while insufficient to protect human
16
17 7 and environmental health, have stimulated extensive innovations in industry, creating jobs while
18
19 8 improving human and environmental health (Wallace 1995, Pearce and Stillwell 2008). These
20
21 9 models can be significantly improved using initiatives such as cradle-to-cradle manufacturing
22
23 10 (Braungart et al. 2007), the bio-economy (McCormick and Kautto 2013), and the increasingly
24
25 11 loud call for a “Green New Deal” (Jones 2009). Embracing such transitions will support our
26
27 12 rights to a pure and high-quality environment. Economically, it will reduce and eventually
28
29 13 eliminate compliance costs, increase labor productivity, provide greater long-run certainty over
30
31 14 operational costs, reduce the economic burden of healthcare costs on society, and increase the
32
33 15 economic advantage of US industries (Braungart et al. 2007, Jones 2009). While some polluting
34
35 16 industries may oppose such initiatives, the above arguments invalidate their rhetorical claims
36
37 17 about the need to reduce regulations to protect jobs and economic advantage. In the face of such
38
39 18 path dependency, it has become incumbent upon the scientific community to evolve industries to
40
41 19 eliminate harms from toxic chemicals, especially given their role in accelerating the existential
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43 20 threat of rapid anthropogenic climate change.
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52 22 **Concluding Remarks**
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3 1 After over 40 years of modern environmental regulations, toxic chemical risks remain
4
5 2 pervasive and largely unacknowledged in the USA, despite their significant negative impacts on
6
7 3 public health, the economy, and life-sustaining ecosystems. Crisis response and risk mitigation
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9 4 have pervaded environmental regulations around toxic chemicals. Effective toxic chemical
10
11 5 governance will require sustained effort to produce better knowledge in the service of large-scale
12
13 6 industrial and social transformations and the creation of inclusive governance bodies. A
14
15 7 transition to a regenerative economy that eliminates the concept of waste and permissible harm is
16
17 8 urgently needed. To do so, researchers, industries, communities, policy makers, and the media
18
19 9 must continue to craft collaborative visions and produce knowledge that enable public and
20
21 10 private investments in clean and ecologically sound technologies and land management
22
23 11 practices. Evaluation of existing systems highlights research priorities for those seeking to
24
25 12 transform governance to improve human and environmental health. By lifting the veil around the
26
27 13 science and technology of producing and managing toxic chemical exposure risks, we can
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29 14 improve democratic governance and insure a healthier, economically robust, and equitable future
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31 15 for all.
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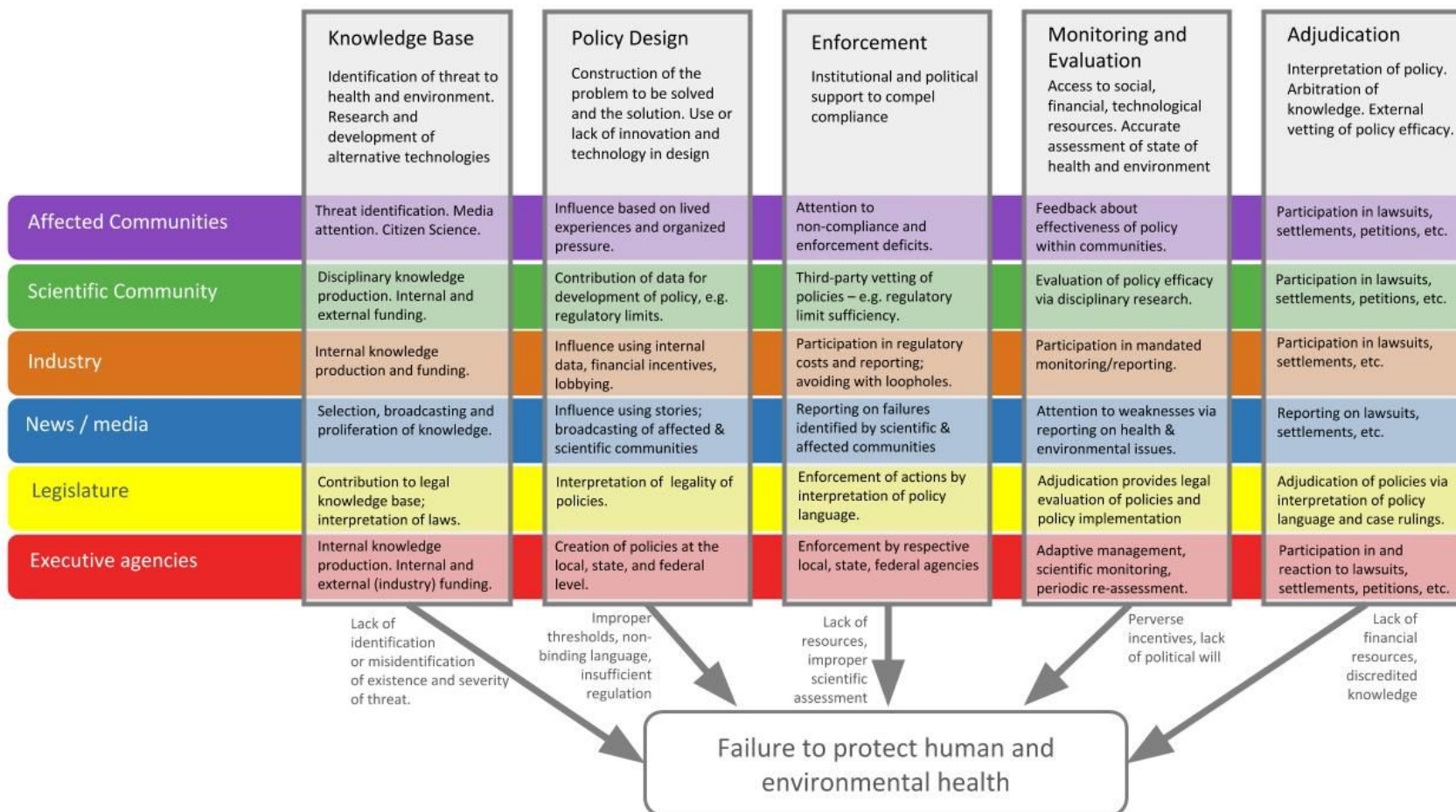


Figure 1. Conceptual framework of U.S. toxics governance domains, and the actors who contribute to each domain. At the base of each box, a summary of common ways these domains contribute toward governance failures are listed. Affected communities bear the costs and risks of contaminants, or benefit from current practices. The scientific community can include academic, government, and

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3 industry scientists, who identify and define risk, but also develop the technologies that create risks. Industry refers to private
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5 companies creating and owning the technologies producing and/or using chemicals. News/media outlets procure and disseminate
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	KNOWLEDGE PRODUCTION	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Nonexistent	No addressing of risks	Nonexistent	No evaluation of policy effectiveness	No independent judiciary
	Constrained scope, widespread disagreement	Disagreement about adequacy (e.g. known loopholes), risks shifted to marginalized communities	Uneven across communities, weak, delayed, or underfunded	Partisan or fragmented, and/or not based off best available information	Court system biased towards specific interests and knowledge when interpreting law and events
	Robust, plural, and in agreement about adequacy	Works to eliminate risk in all communities, transparent distribution of costs and benefits	Transparently utilizes independent monitoring, corrective and punitive actions are swift	Plural and robust framings of causality, and changes in the social distribution of risks, costs, and benefits	Transparent procedures insure accountability for violations of spirit of regulations, designate liability appropriately
EXPERTISE	No knowledge considered authoritative	Ignores most relevant experiences and expertise	Nonexistent	No agreement on feasibility of policy evaluation	Adjudication relies only on legal expertise
	Legitimacy limited to industry or scientific professionals; produced opaquely	Disagreement over utilization of only certain forms of knowledge/expertise	Disagreement about adequacy of M&E processes	Disagreement over adequacy of process for evaluating policy	Adjudication processes allow only certain types of expert knowledge to be considered
	Widespread and plural agreement on transparency, sources are varied, equally scrutinized, and appropriately utilized	Policy process equally weighs diverse forms of expertise; adequate funding for knowledge built in	Widespread agreement that M&E processes are adequate and robust, including localized and decentralized accounts of toxic failures	Includes a range of expertise, utilizes industry-specific protocols, and identifies systemic connections to inform policy changes	Recognizes multiple forms of expertise. Experiential harms are valued; knowledge from adjudication process informs other domains
PATH DEPENDENCY	Inadequate and stagnant	Based exclusively on policy history	Nonexistent or prohibitively underfunded	Nonexistent or not applicable to evolving design	Adjudication increases disagreement and confusion about processes under consideration
	Passively shifts according to technological change without directionality	Minor and iterative changes due to inadequate capacity or oppositional politics	Does not keep pace with changes in knowledge, technology, or context	Underfunded, opaque, and/or privileges the status quo	Disagreement over fairness, relief is partial and does not address governance challenges
	Information and technologies and evaluates its social and technological accomplishments and trajectory	Policy architecture maintains a transparent incentive structure, and yet is sustainable and adaptive to changing circumstances and new knowledge	Independence from regulated entities via sustained, independent funding. Procedures adaptive to technological, social, and environmental change	Adapts to include best practice methods are transparent and accountable; resources are sufficient for adaptive evaluation	Adequate updating and contextualizing precedence and liability

Figure 2: Evaluative framework for examining governance issues of risk, expertise, and path dependency in each governance domain. Green highlighting indicates success (agreement), grey indicates partial success (partial agreement), and red indicates failure (disagreement).

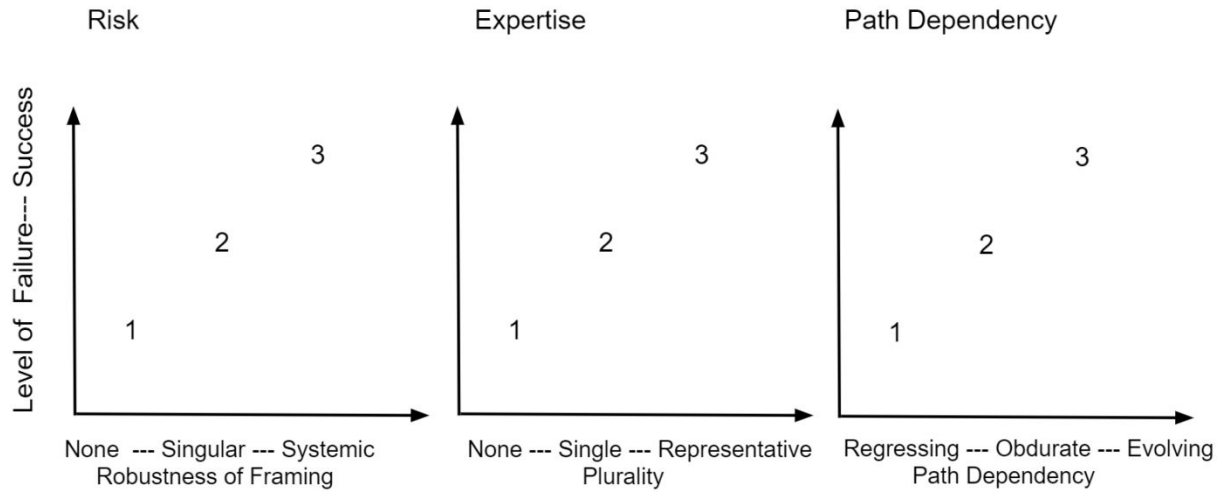


Figure 3: Each indicator was evaluated by level of failure and success (y-axis) and criteria based on attributes of each indicator (x-axis).

A. Case Study: Lead in school drinking water					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	The health risks of soluble and particulate lead, the social and racial contexts of lead exposure, and the role of water treatment operations.	Risks primarily defined as exposure to soluble lead concentrations, disregarding concerns around allowable lead concentrations in plumbing under the Safe Drinking Water Act, as well as infrastructure operation and particulate issues.	Procedures do not adequately assess risk; SDWA monitoring and enforcement only pertains to water supply, not monitoring at tap. Voluntary monitoring does not take into account particulate risks.	Nonexistent at federal level. Local initiatives in relative infancy, though appear to increase engagement of affected communities and have eliminated some systemic drivers of exposure.	No clear assignment of responsibility despite extensive litigation. Remediation often relegated to water filters, with unclear responsibilities for maintenance schedules.
EXPERTISE	Lead may be at harmful levels in drinking water, but community experiences have been disregarded and knowledge collection remains opaque.	Physical science integration remains limited (e.g. particulate issues, dishonest lead-free certification), integrating social dimensions remains largely an unfunded environmental justice mandate at EPA.	Arguments around what constitutes relevant monitoring protocols; extensive work on unequal spatial and social distribution of lead in school drinking water risks not translated into adequate monitoring regimes.	Academic scientists and others outside of formal policy making have conducted evaluations of the overall policy framework, but such evaluations remain external to policy processes.	Courts have called for new information, but significant lack of consensus reflects political economy of interests.
PATH DEPENDENCY	Knowledge production remains focused on increasing precision of estimates of exposure instead of producing relevant knowledge about the systematic and lived risks of affected communities.	The issue of lead risk in school drinking water has been known for some time with no political motivation or resources allocated to create robust federal policy.	Despite widespread knowledge of inadequacy of current monitoring and enforcement regime, limited incentives exist to change behavior by responsible agencies.	Unresolved issues with monitoring prevent effective evaluation in vicious cycle, policy responses remain complaint driven.	Adjudication burden falls to citizens. Policy loopholes and regulatory hierarchy make it difficult to assign liability. Voluntary programs emerging from litigation may show promise for evolving national scale policy framework.

B. Case Study: Heavy metals in light industry					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Health risks of airborne heavy metals are well documented and understood by scientists, government agencies, and the public, including their social distribution.	While policy is designed to mitigate risk from large-scale operations, it misses the risks of small operations, thus shifting the burden to those in range of small-scale industries operating under the loophole.	Heavy metals release exceeded EPA limits, yet because of inadequate regulations, glass companies were in compliance with inadequate permits. State resources were not allocated to monitor the risks of small operations.	National policies have not been re-evaluated, but Oregon regulations for glass furnaces have changed in response to the moss study, and Cleanr Air Oregon, which closes federal regulatory gaps, was adopted.	Proper adjudication was not conducted prior to discovery of environmental failure. Citizens are unsatisfied with the outcome and have taken to litigation, with locally successgful outcomes.
EXPERTISE	Oregon Department of Environmental Quality, USFS, and local citizens had knowledge of heavy metal emissions contamination and had reported it, but no action was taken and complaints were left unresolved.	EPA regulatory loophole exempted small scale glass manufacturing and allowed for privileging of industry knowledge while citizen concerns were ignored.	Formal monitoring by agencies was not conducted. Community members reported concerns of contaminant release, but actions were not taken until agency scientists unintentionally discovered high levels of heavy metals.	Cleaner Air Oregon was developed as a result of collaboration between regulators, citizens, industry, and scientists. However, the exemptions that led to this case study still exist at the national level under the Clean Air Act.	Adjudication was not conducted prior to discovery of environmental failure, thus ignored broader expertise of local residents.
PATH DEPENDENCY	Community complaints and social risks of contaminant exposure continue to be a low priority in the acknowledgement of risk. Manufacturing sector technologies remain dated with limited research for innovation change.	EPA regulatory loopholes are based exclusively on policy history: exemptions exist for small scale industry despite their potential risk to the community, and focus is more on manufacturer-level economic concerns than risks to society.	Toxic release of heavy metals by small scale industry would have continued if not for unintentional detection. Community concerns did not gain power until agencies' scientific findings supported their concerns.	No policy evaluation was conducted until scientific data and community pressure demonstrated a potential human health risk. Now, policies have been improved in Oregon, but not nationally.	EPA has been recently ordered to update its risk reviews for 9 criteria pollutants, but these actions remain overdue and are yet unproven to mitigate for risk.

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C. Case Study: Sulphur and Nitric Oxides					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	There is a substantial body of knowledge of the impacts of SO _x and NO _x on human health and natural environment (although exact effects are uncertain)	The Acid Rain Program established a cap-and-trade program to reduce major emissions of SO _x and NO _x from power plants and create tighter standards on vehicle emissions. However, many rules remain administrative and subject to rollback by executive decision-making, and limited agreement exists on how policy should address risks.	Most power generating facilities have Continuous Emissions Monitors reporting smokestack pollution into an emissions trading market with penalties for noncompliance. Vehicle emissions however remain unmonitored, reliant on inaccurate testing protocols, and emissions standards have been systematically avoided by major automobile manufacturers.	SO _x and NO _x emissions have significantly declined at the national level, but the slow reversal of harms of acid rain on ecosystems remain contingent upon complex factors. Catalytic converter technology has shifted risks from vehicle emissions from SO _x and NO _x to ground level ozone, which has increased due to rising automobile densities.	Successful lawsuits against American Electric Power Co., VW Group North America, and General Motors for failing to meet emissions standards illustrate that the risks of SO _x and NO _x are agreed upon in litigation.
EXPERTISE	A diverse set of disciplinary knowledge created a fairly comprehensive understanding of the sources and impacts of SO _x on human and environmental health. However, industry sectors are the primary knowledge producers. Strategies for reducing emissions from urban and regional planning are not well integrated.	Policy framing remains based around technological and market-based solutions, while systemic solutions framed by academics are less likely to be included in policy discourse.	Community level impacts of pollution, and the social unevenness of them, poorly considered in monitoring. Because monitoring occurs at the pipe, knowledge produced about ambient levels within urban and other affected communities is ignored, except in extreme cases.	End of pipe expertise remains dominant in policy evaluation; little consideration is given to the systemic and risk-shifting aspects of both stationary and mobile sources.	Attempts to broaden the mandate of the EPA to promote systemic energy and transportation transitions have been stalled by the Supreme Court. The EPA under new administration has reversed its position and has reverted to seeking input on appropriate regulation only from the regulated industries.
PATH DEPENDENCY	There have been significant investments in developing technologies of 'scrubbing' SO _x and NO _x from fossil fuel combustion sources, but overall reliance on fossil fuel-based power generation, heating, and transportation remains largely unchanged	Disagreement persists about the need for comprehensive overhaul and guidance as to the complex social and technological drivers of emissions. However, federal and state energy policies, and energy markets, appear to be shifting the grid generation mix away from emissions heavy sources of coal and oil and towards renewables and natural gas	Enforcement regimes depend on administrative policy architecture. Current monitoring regime insufficient to parse liability for fluctuations in regional air quality due to mobile sources.	While recent initiatives have expanded regulatory oversight of SO _x and NO _x emissions, present administrative changes threaten to reverse the systemic overhaul of the power production sector via the Clean Power Plan. Difficulties with reliably enforcing vehicle emissions standards. Limited push to evolve transportation policy away from fossil fuel dependency.	Doctrines of liability in USA, combined with unlimited corporate influence in elections, have compromised independence of judiciary and enshrined private property rights at cost of public health.

D. Case Study: Bisphenol-a (BPA)

	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Risk was not adequately evaluated prior to market release. It is now well accepted by the public and scientific community that BPA is unsafe to use in food and beverage containers, especially for babies and young children, however the perceived risk is mismatched between industry, and scientists and the public.	Under TSCA, compounds are not required to be tested and evaluated for their chronic long-term effects, biophysical risks, or societal risks before going to market, as burden of proof is placed on industry labs or industry-contracted scientists. Post production burden of proof for harm is placed on victims.	There is no formal process for monitoring the risks of compounds under TSCA, and any monitoring that does exist is inadequate for substances that have cumulative risk factors and are not acute toxic substances.	Formal policy evaluation not required, and existing evaluations are reactionary. Changes due to evaluation are incremental and do not typically lead to changes in policy design or reassessment of risk. Decreasing use is due to market pressure, which has primarily focused on labeling, not chemistry of replacement substances which may be just as toxic.	Petition filed to the FDA resulted in the amendment of food additive regulations to no longer "provide for the use of BPA" in baby bottles and sippy cups, and infant formula packaging. This change only narrowly addresses the full array of risks associated with BPA and does not assign liability to manufacturers.
EXPERTISE	Knowledge is primarily produced using industry standard dose-response methods instead of research on of endocrine disruption caused by environmentally relevant concentrations.	TSCA and related policies overseeing chemicals like BPA do not adequately consult multiple forms of expertise. Data (or lack thereof) produced by industry-sponsored labs is the primary determinant of market release and/or restrictions on use.	Monitoring of long-term effects of chemicals on the market, such as BPA, is not required of manufacturers or enforcement agencies. Instead, scientists and communities independently serve as "monitors."	TSCA was updated in 2016 due to significant public and scientific pressure. Pre-market review of chemicals added, but disagreement remains over adequacy; evaluation of overall impact on protecting human and environmental health limited.	The passing of the above amendments shows that adjudication can be successful if it requires minor actions that take minimal enforcement, but does not address risks produced by all forms of knowledge.
PATH DEPENDENCY	Manufacturing practices have started to acknowledge the risks of BPA, however, the systematic way in which knowledge is produced for governance has remained constant; BPA merely replaced by new chemicals that are not adequately tested.	Current FDA policies now no longer list BPA as an approved additive in certain products, but this is less a recognition of its health hazard and more a reaction to its discontinued use by industry due to consumer pressure.	Monitoring is not required by TSCA.	Policy evaluation does not occur unless significant evidence of toxic risk and significant pressure exists, in which case incremental changes occur that tend to perpetuate status quo, as seen by de-listing BPA as an additive instead of banning it or requiring re-evaluation.	Adjudication has been mildly successful, but has come at the expense of those who petitioned the FDA. Disagreement remains on the effectiveness of adjudication in resolving the chemical's underlying risk; outcomes have not substantially improved governance of chemicals like BPA.

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E. Case Study: Glyphosate in agriculture					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Health risk contested; health standards based on a dose-response approach, despite evidence of endocrine disruption at environmentally relevant concentrations. Some studies report carcinogenic and endocrine-disrupting properties, while EPA testing has not.	The Food Quality Protection Act allows industry to be the primary assessors of consumer risk, and Worker Protection Standards only pertain to applicators. Both policies shift burden of risk to consumers, agricultural communities, and ecosystems. Federal re-certification requires periodic re-assessment every 10 years.	No formal mechanisms for assessing persistent and cumulative risk of glyphosate use. The Federal Insecticide, Fungicide, and Rodenticide Act (FIFRA) requires commercial users to hold an applicators license, and pesticides to have proper labeling and directions, but applicators are not required to report usage.	Extensive evaluation has been conducted by the scientific community demonstrating the need for a precautionary approach, however, industry-funded studies and lobbying hampers the actions necessary to mitigate risk; despite existing regulations, evidence indicates risks of pesticides have continued to increase.	Definitions of risk highly contested, historically relying on proof of acute toxicity and industry generated studies (e.g. within CA Prop. 65 listing, pending FIFRA re-certification). Current court cases have opened up risk characterization and issues has received increased attention by media, scientific community, and affected communities.
EXPERTISE	Traditional dose-response approaches privileged over others such as studies of endocrine disruption and epidemiological approaches, leading to increased application level allowances, despite growing evidence of health and environmental danger.	Lack of precautionary measures and are driven by industry lobbying. Does not consider peer-reviewed literature or affected communities' concerns, except during re-certification. Consistently disregards knowledge on alternative means of crop protection.	No formal requirements for monitoring, independent scientific research is typically the only source of data, often in response to community concerns and/or pressure, but the results are not linked to enforcement protocols and have limited ability to change standards.	Policy evaluation continues to examine pesticide policies based on needs and concerns of producing industries, not the social and environmental knowledge base.	Historically did not consider broader expertise, and burden of proof linking specific health and environmental impacts falls on affected communities and public interest scientists in face of extensive industry sponsored research.
PATH DEPENDENCY	Food system has become increasingly dependent on increasing inputs of glyphosate and other pesticides of known toxicity, despite long term warnings about the need for holistic agricultural management. Agencies appear to favor industry-produced knowledge and risk assessment.	Continues to disproportionately benefit industrial operations reliant on increasing levels of chemical additives, only regulated to prevent acute harm. Recertification processes provide limited adaptive governance potential.	Formal monitoring and enforcement activities are not required by pesticide policies and results of independent monitoring activities have limited ability to change pesticide standards.	Profound disagreements persist over necessity and scope of evaluation.	Compensatory wins' possible for affected communities and individuals, have come at great cost, and show limited potential to shift overall industry model of producing toxic chemicals for agricultural production.

36 Figure 4. Application of indicators to five case studies across multiple chemical classes, including (A) Lead in school
37 drinking water, (B) Heavy metal emissions from light industry, (C) Sulphur and nitric oxide emissions, (D) Bisphenol-a,
38 and (E) Glyphosate use in agriculture. For references, see Supplemental Information 1.
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5 **Case Study E: Glyphosate use in agriculture.**
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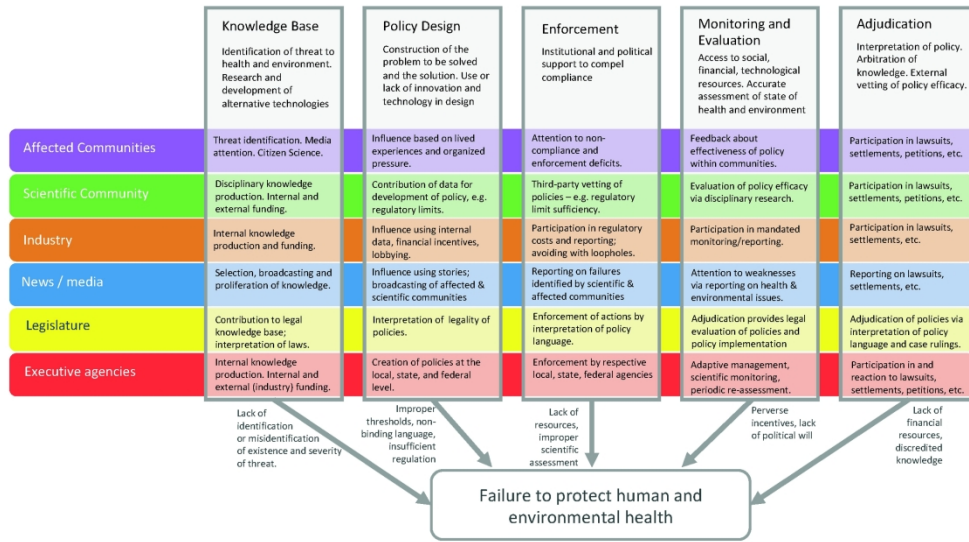
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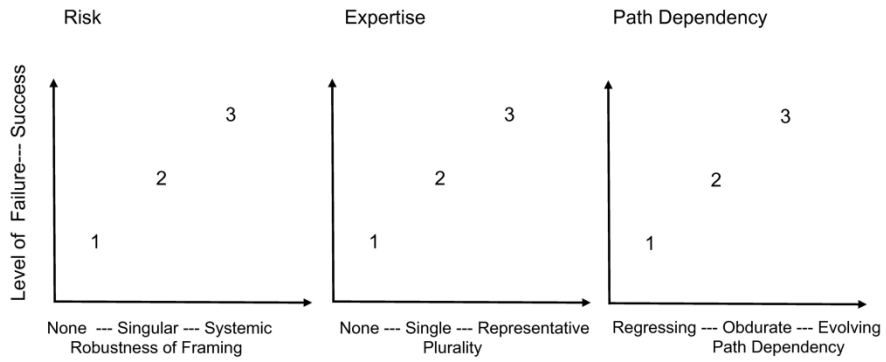
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	KNOWLEDGE PRODUCTION	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Nonexistent	No addressing of risks	Nonexistent	No evaluation of policy effectiveness	No independent judiciary
	Constrained scope, widespread disagreement	Disagreement about adequacy (e.g. known loopholes), risks shifted to marginalized communities	Uneven across communities, weak, delayed, or underfunded	Partisan or fragmented, and/or not based off best available information	Court system biased towards specific interests and knowledge when interpreting law and events
	Robust, plural, and in agreement about adequacy	Works to eliminate risk in all communities, transparent distribution of costs and benefits	Transparently utilizes independent monitoring, corrective and punitive actions are swift	Plural and robust framings of causality, and changes in the social distribution of risks, costs, and benefits	Transparent procedures insure accountability for violations of spirit of regulations, designate liability appropriately
EXPERTISE	No knowledge considered authoritative	Ignores most relevant experiences and expertise	Nonexistent	No agreement on feasibility of policy evaluation	Adjudication relies only on legal expertise
	Legitimacy limited to industry or scientific professionals; produced opaquely	Disagreement over utilization of only certain forms of knowledge/expertise	Disagreement about adequacy of M&E processes	Disagreement over adequacy of process for evaluating policy	Adjudication processes allow only certain types of expert knowledge to be considered
	Widespread and plural agreement on transparency, sources are varied, equally scrutinized, and appropriately utilized	Policy process equally weighs diverse forms of expertise; adequate funding for knowledge built in	Widespread agreement that M&E processes are adequate and robust, including localized and decentralized accounts of toxic failures	Includes a range of expertise, utilizes industry-specific protocols, and identifies systemic connections to inform policy changes	Recognizes multiple forms of expertise. Experiential harms are valued, knowledge from adjudication process informs other domains
PATH DEPENDENCY	Inadequate and stagnant	Based exclusively on policy history	Nonexistent or prohibitively underfunded	Nonexistent or not applicable to evolving design	Adjudication increases disagreement and confusion about processes under consideration
	Passively shifts according to technological change without directionality	Minor and iterative changes due to inadequate capacity or oppositional politics	Does not keep pace with changes in knowledge, technology, or context	Underfunded, opaque, and/or privileges the status quo	Disagreement over fairness, relief is partial and does not address governance challenges
	Information and technologies and evaluates its social and technological accomplishments and trajectory	Policy architecture maintains a transparent incentive structure, and yet is sustainable and adaptive to changing circumstances and new knowledge	Independence from regulated entities via sustained, independent funding. Procedures adaptive to technological, social, and environmental change	Adapts to include best practice methods are transparent and accountable; resources are sufficient for adaptive evaluation	Adequate updating and contextualizing precedence and liability

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A. Case Study: Lead in school drinking water					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	The health risks of soluble and particulate lead, the social and racial contexts of lead exposure, and the role of water treatment operations.	Risks primarily defined as exposure to soluble lead concentrations, disregarding concerns around allowable lead concentrations in plumbing under the Safe Drinking Water Act, as well as infrastructure operation and particulate issues.	Procedures do not adequately assess risk; SDWA monitoring and enforcement only pertains to water supply, not monitoring at tap. Voluntary monitoring does not take into account particulate risks.	Nonexistent at federal level. Local initiatives in relative infancy, though appear to increase engagement of affected communities and have eliminated some systemic drivers of exposure.	No clear assignment of responsibility despite extensive litigation. Remediation often relegated to water filters, with unclear responsibilities for maintenance schedules.
EXPERTISE	Lead may be at harmful levels in drinking water, but community experiences have been disregarded and knowledge collection remains opaque.	Physical science integration remains limited (e.g. particulate issues, dishonest lead-free certification), integrating social dimensions remains largely an unfunded environmental justice mandate at EPA.	Arguments around what constitutes relevant monitoring protocols; extensive work on unequal spatial and social distribution of lead in school drinking water risks not translated into adequate monitoring regimes.	Academic scientists and others outside of formal policy making have conducted evaluations of the overall policy framework, but such evaluations remain external to policy processes.	Courts have called for new information, but significant lack of consensus reflects political economy of interests.
PATH DEPENDENCY	Knowledge production remains focused on increasing precision of estimates of exposure instead of producing relevant knowledge about the systematic and lived risks of affected communities.	The issue of lead risk in school drinking water has been known for some time with no political motivation or resources allocated to create robust federal policy.	Despite widespread knowledge of inadequacy of current monitoring and enforcement regime, limited incentives exist to change behavior by responsible agencies.	Unresolved issues with monitoring prevent effective evaluation in vicious cycle, policy responses remain complaint driven.	Adjudication burden falls to citizens. Policy loopholes and regulatory hierarchy make it difficult to assign liability. Voluntary programs emerging from litigation may show promise for evolving national scale policy framework.

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B. Case Study: Heavy metals in light industry					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Health risks of airborne heavy metals are well documented and understood by scientists, government agencies, and the public, including their social distribution.	While policy is designed to mitigate risk from large-scale operations, it misses the risks of small operations, thus shifting the burden to those in range of small-scale industries operating under the loophole.	Heavy metals release exceeded EPA limits, yet because of inadequate regulations, glass companies were in compliance with inadequate permits. State resources were not allocated to monitor the risks of small operations.	National policies have not been re-evaluated, but Oregon regulations for glass furnaces have changed in response to the moss study, and Cleanr Air Oregon, which closes federal regulatory gaps, was adopted.	Proper adjudication was not conducted prior to discovery of environmental failure. Citizens are unsatisfied with the outcome and have taken to litigation, with locally successful outcomes.
EXPERISE	Oregon Department of Environmental Quality, USFS, and local citizens had knowledge of heavy metal emissions contamination and had reported it, but no action was taken and complaints were left unresolved.	EPA regulatory loophole exempted small scale glass manufacturing and allowed for privileging of industry knowledge while citizen concerns were ignored.	Formal monitoring by agencies was not conducted. Community members reported concerns of contaminant release, but actions were not taken until agency scientists unintentionally discovered high levels of heavy metals.	Cleaner Air Oregon was developed as a result of collaboration between regulators, citizens, industry, and scientists. However, the exemptions that led to this case study still exist at the national level under the Clean Air Act.	Adjudication was not conducted prior to discovery of environmental failure, thus ignored broader expertise of local residents.
PATH DEPENDENCY	Community complaints and social risks of contaminant exposure continue to be a low priority in the acknowledgement of risk. Manufacturing sector technologies remain dated with limited research for innovation change.	EPA regulatory loopholes are based exclusively on policy history: exemptions exist for small scale industry despite their potential risk to the community, and focus is more on manufacturer-level economic concerns than risks to society.	Toxic release of heavy metals by small scale industry would have continued if not for unintentional detection. Community concerns did not gain power until agencies' scientific findings supported their concerns.	No policy evaluation was conducted until scientific data and community pressure demonstrated a potential human health risk. Now, policies have been improved in Oregon, but not nationally.	EPA has been recently ordered to update its risk reviews for 9 criteria pollutants, but these actions remain overdue and are yet unproven to mitigate for risk.

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C. Case Study: Sulphur and Nitric Oxides					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	There is a substantial body of knowledge of the impacts of SO _x and NO _x on human health and natural environment (although exact effects are uncertain)	The Acid Rain Program established a cap-and-trade program to reduce major emissions of SO _x and NO _x from power plants and create tighter standards on vehicle emissions. However, many rules remain administrative and subject to rollback by executive decision-making, and limited agreement exists on how policy should address risks.	Most power generating facilities have Continuous Emissions Monitors reporting smoketack pollution into an emissions trading market with penalties for noncompliance. Vehicle emissions however remain unmonitored, reliant on inaccurate testing protocols, and emissions standards have been systematically avoided by major automobile manufacturers.	SO _x and NO _x emissions have significantly declined at the national level, but the slow reversal of harms of acid rain on ecosystems remain contingent upon complex factors. Catalytic converter technology has shifted risks from vehicle emissions from SO _x and NO _x to ground level ozone, which has increased due to rising automobile densities.	Successful lawsuits against American Electric Power Co., VW Group North America, and General Motors for failing to meet emissions standards illustrate that the risks of SO _x and NO _x are agreed upon in litigation.
EXPERTISE	A diverse set of disciplinary knowledge created a fairly comprehensive understanding of the sources and impacts of SO _x on human and environmental health. However, industry sectors are the primary knowledge producers. Strategies for reducing emissions from urban and regional planning are not well integrated.	Policy framing remains based around technological and market-based solutions, while systemic solutions framed by academics are less likely to be included in policy discourse.	Community level impacts of pollution, and the social unevenness of them, poorly considered in monitoring. Because monitoring occurs at the pipe, knowledge produced about ambient levels within urban and other affected communities is ignored, except in extreme cases.	End of pipe expertise remains dominant in policy evaluation; little consideration is given to the systemic and risk-shifting aspects of both stationary and mobile sources.	Attempts to broaden the mandate of the EPA to promote systemic energy and transportation transitions have been stalled by the Supreme Court. The EPA under new administration has reversed its position and has reverted to seeking input on appropriate regulation only from the regulated industries.
PATH DEPENDENCY	There have been significant investments in developing technologies of 'scrubbing' SO _x and NO _x from fossil fuel combustion sources, but overall reliance on fossil fuel-based power generation, heating, and transportation remains largely unchanged	Disagreement persists as to the need for comprehensive overhaul and guidance as to the complex social and technological drivers of emissions. However, federal and state energy policies, and energy markets, appear to be shifting the grid generation mix away from emissions heavy sources of coal and oil and towards renewables and natural gas	Enforcement regimes depend on administrative policy architecture. Current monitoring regime insufficient to parse liability for fluctuations in regional air quality due to mobile sources.	While recent initiatives have expanded regulatory oversight of SO _x and NO _x emissions, present administrative changes threaten to reverse the systemic overhaul of the power production sector via the Clean Power Plan. Difficulties with reliably enforcing vehicle emissions standards. Limited push to evolve transportation policy away from fossil fuel dependency.	Doctrines of liability in USA, combined with unlimited corporate influence in elections, have compromised independence of judiciary and enshrined private property rights at cost of public health.

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D. Case Study: Bisphenol-a (BPA)

	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Risk was not adequately evaluated prior to market release. It is now well accepted by the public and scientific community that BPA is unsafe to use in food and beverage containers, especially for babies and young children, however the perceived risk is mismatched between industry, and scientists and the public.	Under TSCA, compounds are not required to be tested and evaluated for their chronic long-term effects, biophysical risks, or societal risks before going to market, as burden of proof is placed on industry labs or industry-contracted scientists. Post production burden of proof for harm is placed on victims.	There is no formal process for monitoring the risks of compounds under TSCA, and any monitoring that does exist is inadequate for substances that have cumulative risk factors and are not acute toxic substances.	Formal policy evaluation not required, and existing evaluations are reactionary. Changes due to evaluation are incremental and do not typically lead to changes in policy design or reassessment of risk. Decreasing use is due to market pressure, which has primarily focused on labeling, not chemistry of replacement substances which may be just as toxic.	Petition filed to the FDA resulted in the amendment of food additive regulations to no longer "provide for the use of BPA" in baby bottles and sippy cups, and infant formula packaging. This change only narrowly addresses the full array of risks associated with BPA and does not assign liability to manufacturers.
EXPERTISE	Knowledge is primarily produced using industry standard dose-response methods instead of research on of endocrine disruption caused by environmentally relevant concentrations.	TSCA and related policies overseeing chemicals like BPA do not adequately consult multiple forms of expertise. Data or lack thereof produced by industry-sponsored labs is the primary determinant of market release and/or restrictions on use.	Monitoring of long-term effects of chemicals on the market, such as BPA, is not required of manufacturers or enforcement agencies. Instead, scientists and communities independently serve as "monitors."	TSCA was updated in 2016 due to significant public and scientific pressure. Pre-market review of chemicals added, but disagreement remains over adequacy; evaluation of overall impact on protecting human and environmental health limited.	The passing of the above amendments shows that adjudication can be successful if it requires minor actions that take minimal enforcement, but does not address risks produced by all forms of knowledge.
PATH DEPENDENCY	Manufacturing practices have started to acknowledge the risks of BPA, however, the systematic way in which knowledge is produced for governance has remained constant; BPA merely replaced by new chemicals that are not adequately tested.	Current FDA policies now no longer list BPA as an approved additive in certain products, but this is less a recognition of its health hazard and more a reaction to its discontinued use by industry due to consumer pressure.	Monitoring is not required by TSCA.	Policy evaluation does not occur unless significant evidence of toxic risk and significant pressure exists, in which case incremental changes occur that tend to perpetuate status quo, as seen by de-listing BPA as an additive instead of banning it or requiring re-evaluation.	Adjudication has been mildly successful, but has come at the expense of those who petitioned the FDA. Disagreement remains on the effectiveness of adjudication in resolving the chemical's underlying risk; outcomes have not substantially improved governance of chemicals like BPA.

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E. Case Study: Glyphosate in agriculture					
	KNOWLEDGE BASE	POLICY DESIGN	MONITORING AND ENFORCEMENT	POLICY EVALUATION	ADJUDICATION
RISK	Health risk contested; health standards based on a dose-response approach, despite evidence of endocrine disruption at environmentally relevant concentrations. Some studies report carcinogenic and endocrine-disrupting properties, while EPA testing has not.	The Food Quality Protection Act allows industry to be the primary assessors of consumer risk, and Worker Protection Standards only pertain to applicators. Both policies shift burden of risk to consumers, agricultural communities, and ecosystems. Federal re-certification requires periodic re-assessment every 10 years.	No formal mechanisms for assessing persistent and cumulative risk of glyphosate use. The Federal Insecticide, Fungicide, and Rodenticide Act (FIFRA) requires commercial users to hold an applicators license, and pesticides to have proper labeling and directions, but applicators are not required to report usage.	Extensive evaluation has been conducted by the scientific community demonstrating the need for a precautionary approach, however, industry-funded studies and lobbying hampers the actions necessary to mitigate risk; despite existing regulations, evidence indicates risks of pesticides have continued to increase.	Definitions of risk highly contested, historically relying on proof of acute toxicity and industry generated studies (e.g. within CA Prop. 65 listing, pending FIFRA re-certification). Current court cases have opened up risk characterization and issues has received increased attention by media, scientific community, and affected communities.
EXPERTISE	Traditional dose-response approaches privileged over others such as studies of endocrine disruption and epidemiological approaches, leading to increased application level allowances, despite growing evidence of health and environmental danger.	Lack of precautionary measures and are driven by industry lobbying. Does not consider peer-reviewed literature or affected communities' concerns, except during re-certification. Consistently disregards knowledge on alternative means of crop protection.	No formal requirements for monitoring, independent scientific research is typically the only source of data, often in response to community concerns and/or pressure, but the results re not linked to enforcement protocols and have limited ability to change standards.	Policy evaluation continues to examine pesticide policies based on needs and concerns of producing industries, not the social and environmental knowledge base.	Historically did not consider broader expertise, and burden of proof linking specific health and environmental impacts falls on affected communities and public interest scientists in face of extensive industry sponsored research.
PATH DEPENDENCY	Food system has become increasingly dependent on increasing inputs of glyphosate and other pesticides of known toxicity, despite long term warnings about the need for holistic agricultural management. Agencies appear to favor industry-produced knowledge and risk assessment.	Continues to disproportionately benefit industrial operations reliant on increasing levels of chemical additives, only regulated to prevent acute harm. Recertification processes provide limited adaptive governance potential.	Formal monitoring and enforcement activities are not required by pesticide policies and results of independent monitoring activities have limited ability to change pesticide standards.	Profound disagreements persist over necessity and scope of evaluation.	Compensatory wins' possible for affected communities and individuals, have come at great cost, and show limited potential to shift overall industry model of producing toxic chemicals for agricultural production.

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