Articles

# Freshwater Ecoregions of the World: A New Map of Biogeographic Units for Freshwater Biodiversity Conservation

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We present a new map depicting the first global biogeographic regionalization of Earth's freshwater systems. This map of freshwater ecoregions is based on the distributions and compositions of freshwater fish species and incorporates major ecological and evolutionary patterns. Covering virtually all freshwater habitats on Earth, this ecoregion map, together with associated species data, is a useful tool for underpinning global and regional conservation planning efforts (particularly to identify outstanding and imperiled freshwater systems); for serving as a logical framework for large-scale conservation strategies; and for providing a global-scale knowledge base for increasing freshwater biogeographic literacy. Preliminary data for fish species compiled by ecoregion reveal some previously unrecognized areas of high biodiversity, highlighting the benefit of looking at the world's freshwaters through a new framework.

Keywords: freshwater, ecoregions, biogeography, fish, mapping

Growth of the human population, rising consumption, and rapid globalization have caused widespread degradation and disruption of natural systems, especially in the freshwater realm. Freshwater ecosystems have lost a greater proportion of their species and habitat than ecosystems on land or in the oceans, and they face increasing threats from dams, water withdrawals, pollution, invasive species, and overharvesting (MEA 2005, Revenga et al. 2005). Freshwater

ecosystems and the diverse communities of species found in lakes, rivers, and wetlands may be the most endangered of all (MEA 2005).

These stressed systems support an extraordinarily high proportion of the world's biodiversity. In terms of area, freshwater ecosystems occupy only 0.8% of Earth's surface, but they are estimated to harbor at least 100,000 species, or nearly 6% of all described species (Dudgeon et al. 2006). Each year,

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Despite this combination of extraordinary richness, high endemism, and exceptional threat, few broadscale conservation planning efforts have targeted freshwater systems and their dependent species. This relative inattention derives in part from an acute lack of comprehensive, synthesized data on the distributions of freshwater species (Revenga and Kura 2003). The most exhaustive recent global inventory of freshwater taxa acknowledges serious survey gaps and assigns species distributions only to the level of continent (Lévêque et al. 2005). Such inventories are valuable for highlighting research priorities and providing a global picture of how taxonomic diversity compares across continents, but they have limited utility for conservation planning efforts, for which the largest planning unit is often the river basin or ecoregion.

#### A global freshwater regionalization

Ecoregions are a widely recognized and applied geospatial unit for conservation planning, developed to represent the patterns of environmental and ecological variables known to influence the distribution of biodiversity features at broad scales (Groves et al. 2002). Building on the work of Dinerstein and colleagues (1995), we define a freshwater ecoregion as a large area encompassing one or more freshwater systems with a distinct assemblage of natural freshwater communities and species. The freshwater species, dynamics, and environmental conditions within a given ecoregion are more similar to each other than to those of surrounding ecoregions, and together form a conservation unit. Ecoregion boundaries are not necessarily determined by the turnover of species ranges (McDonald et al. 2005) but are intended to describe broad patterns of species composition and associated ecological and evolutionary processes.

Ecoregion delineation benefits from the best available data describing species and systems ecology, but can proceed with imperfect information (Wikramanayake et al. 2002). Global ecoregion frameworks have already been developed for the terrestrial and, more recently, marine realms, both of which are characterized by their own data limitations (Olson et al. 2001, Spalding et al. 2007). In this article we demonstrate how the ecoregion concept has been applied to freshwater systems, and present the first global map of freshwater ecoregions— a starting point for conservation planning anywhere on Earth.

Ecoregions have typically been delineated to represent patterns of potential vegetation (Olson et al. 2001) and have at times been used to characterize regional differences in water quality as well (Omernik 1987). Terrestrial ecoregions are delineated largely on the basis of climate, physiography, and vegetation types, but different features are often dominant in shaping the broadscale distributions of freshwater species. As Tonn (1990) described, the species occurring in a given river reach, lake, spring, or wetland will be a function of a hierarchy of continental-scale filters (including mountain building, speciation, and glaciation) that have defined large biogeographic patterns; regional-scale filters (such as broad climatic and physiographic patterns, and dispersal barriers such as regional catchments); and subregional and finerscale habitat filters (e.g., distinct physiographic types and macrohabitats) acting on the regional species pool. Freshwater ecoregions capture the patterns generated primarily by continental- and regional-scale filters.

Of these filters, dispersal barriers in the form of catchment divides (also called watersheds) are distinctive to freshwaters. Unlike terrestrial species or those with aerial or wind-dispersed life stages, obligate freshwater species-those confined to the freshwater environment and unable to move via land, air, or sea-generally cannot disperse from one unconnected catchment to another. Furthermore, all species dependent on freshwater systems, whether or not they are confined to the aquatic environment, are to some extent affected by the hydrological and linked ecological processes of the catchments where they live. As a result, catchments strongly influence broad freshwater biogeographic patterns in most regions. There are exceptions, however. Tectonic movements have in some cases separated once-joined catchments, allowing for further speciation. Also, natural drainage evolution over geological time includes river piracy, which severs connections and provides new interdrainage links that reconform systems. The freshwater ecoregions of the world presented here reflect both the hydrological underpinning of freshwater fish species distributions as well as historical shifts in landmasses and consequent evolutionary processes.

#### Ecoregion delineation and species list compilation

No global biogeographic framework for freshwater species was available as the foundation for our map. The applicability of Wallace's (1876) and Udvardy's (1975) zoogeographic realms to most freshwater taxa is unresolved (Berra 2001, Vinson and Hawkins 2003), and these divisions are too large for conservation planning endeavors. Several examinations of global freshwater biogeography (e.g., Banarescu 1990) provided information at somewhat finer scales but could not be clearly translated into seamless ecoregion delineations. Where appropriate, we adapted previous continental efforts. For North America, Africa, and Madagascar, we updated regionalizations outlined in two previously published volumes (Abell et al. 2000, Thieme et al. 2005), but we excluded a prior delineation for Latin America and the Caribbean (Olson et al. 1998) because the approach differed markedly from our current methodology, and data have improved substantially since its development (e.g., Reis et al. 2003). We examined but chose to exclude the 25 European regions of Illies's impressive Limnofauna Europaea (1978) because the approach for delineating those regions differed considerably from ours: those regions were based on the distributions of 75 different taxonomic groups and were drawn without reference to catchments. Moreover, neither ecological nor evolutionary processes figured in those delineations. A complete list of all references and experts consulted in the process of delineating ecoregions is available online (*www.feow.org*).

We assembled our global map of freshwater ecoregions using the best available regional information describing freshwater biogeography, defined broadly to include the influences of phylogenetic history, palaeogeography, and ecology (Banarescu 1990). We restricted our analyses to information describing freshwater fish species distributions, with a few exceptions for extremely data-poor regions and inland seas, where some invertebrates and brackish-water fish were considered, respectively. We focused on freshwater fish for several reasons. On a global scale, fish are the best-studied obligate aquatic taxa. Detailed information exists for other freshwater taxa in regions like North America and Europe, but the consideration of such groups in a global analysis would be difficult, given the wide variation in available data (Balian et al. 2008). Freshwater dispersant fish species—those unable to cross saltwater barriers-are better zoogeographic indicators than freshwater invertebrates, which can often disperse over land, survive in humid atmospheres outside water, or be transported between freshwaters (Banarescu 1990). Finally, the distributions of obligate aquatic invertebrate groups in general respond to ecological processes at localized scales that are too small to be meaningful for ecoregion delineation (Wasson et al. 2002). Therefore, fish serve as proxies for the distinctiveness of biotic assemblages. We recognize that analyses of other taxonomic groups would almost certainly reveal different patterns for some regions, and that our results are scale dependent (Paavola et al. 2006). Our near-exclusive focus on fish is a departure from earlier continental ecoregionalization exercises (Abell et al. 2000, Thieme et al. 2005), and we have updated the ecoregion delineations accordingly.

The available data for describing fish biogeography vary widely. In the United States, it is possible to map presence/ absence data for all freshwater fish species to subbasins averaging about 2025 square kilometers (km<sup>2</sup>) in size (NatureServe 2006). But for many of the world's species, occurrence data are limited to a small number of irregularly surveyed systems. Large parts of the massive Congo basin remain unsampled, for instance, with most sampling occurring near major towns and most taxonomic studies of the region dating from the 1960s. Problems with taxonomy and species concepts hamper broadscale analyses even where systems have been reasonably well sampled (Lundberg et al. 2000). Although addressing many of these problems is beyond the scope of this project, in our analyses we have attempted to minimize nomenclatural errors by normalizing species names with Eschmeyer's Catalog of Fishes (2006; www.calacademy.org/ research/ichthyology/catalog/).

Freshwater fish patterns were analyzed separately for different regions of the world to account for data variability. The geographic scope of major information sources largely defined those regions (table 1). Information sources were typically taxonomic works, some of which included biogeographical analyses. Leading ichthyologists delineated ecoregions primarily by examining the distributions of endemic species, genera, and families against the backdrop of an area's dominant habitat features and the presence of ecological (e.g., large concentrations of long-distance migratory species) and evolutionary (e.g., species flocks) phenomena. More than 130 ichthyologists and freshwater biogeographers contributed to the global map by either delineating or reviewing ecoregions.

Data gaps and biogeographic drivers resulted in the use of slightly different criteria among and even within some regions (table 2, box 1). Where fish species data were reasonably comprehensive and available at subbasin or finer scales, we attributed species distributions to catchments to facilitate evaluation of biogeographic patterns in a bottom-up

| Region               | Primary information source   |
|----------------------|--|
| Africa               | Roberts 1975, Skelton 1994, Lévêque 1997, Thieme et al. 2005                       |
| Middle East          | No regional information sources available.   |
| Former USSR          | No regional information sources available.   |
| Remainder of Eurasia | For Europe: Kottelat and Freyhof 2007; no regionwide information sources for Asia. |
| Australasia          | McDowall 1990, Allen 1991, Unmack 2001, Allen et al. 2002                          |
| Oceania              | Keith et al. 2002  |
| Canada               | Scott and Crossman 1998  |
| United States        | Maxwell et al. 1995, Abell et al. 2000   |
| Mexico               | Contreras-Balderas 2000, Miller et al. 2005  |
| Central America      | Bussing 1976, CLOFFSCA (Reis et al. 2003)  |
| Caribbean            | Rauchenberger 1988, Burgess and Franz 1989   |
| South America        | CLOFFSCA (Reis et al. 2003), Menni 2003  |

*Note:* In many cases, these same sources were used to compile species lists. A full bibliography with additional publications, which along with unpublished data often constituted the greater part of inputs to ecoregion delineations and species lists, is available at the Web site *www.feow.org*. Every region also benefited from expert input; individual contributors are listed in the acknowledgments section and at the Web site. Regions in some cases correspond to politically rather than biophysically defined units to take advantage of existing information sources and expertise.

approach. For example, a new high-resolution hydrographic dataset (HydroSHEDS; *www.wwfus.org/freshwater/hydrosheds. cfm*) for South America provided fine-scale catchment maps that, in conjunction with newly synthesized species data (Reis et al. 2003), aided in the assessment of biogeography. In regions without extensive species data, or where major basins support highly similar faunas as a result of recent glaciation, a top-down analysis used qualitative expert knowledge of distinctive species and assemblages to map major biogeographic patterns (table 2). Ecoregional boundaries resulting from either approach, therefore, largely coincide with catchment boundaries.

Whereas overall there is correspondence between catchments and ecoregion boundaries, unconnected neighboring catchments were in some cases grouped together, where strong biogeographic evidence indicates that landscape or other features overrode contemporary hydrographic integrity. For example, owing to historic drainage evolution and similarities in fauna, Africa's southern temperate highveld combines headwaters of coastal basins that drain to the Indian Ocean with those of the Atlantic-draining Orange basin. Considerable faunal exchange of the headwaters of the Orange River system with that of the coastal systems may have occurred as the coastal rivers eroded their basins at a faster rate than the adjacent Orange tributaries (Skelton et al. 1995). These and other examples demonstrate that historical geographic events and current hydrology may have conflicting effects on the fish fauna of a particular region and thereby argue for different boundaries. The decision to weigh some effects more strongly than others was made on a case-by-case basis, and it is acknowledged that additional data may favor alternative delineations.

With the exception of islands, individual freshwater ecoregions typically cover tens of thousands to hundreds of thousands of square kilometers (Maxwell et al. 1995). Ecoregion size varies in large part because of landscape history. Regions with depauperate faunas resulting from recent glaciation events tend to have large ecoregion sizes, as do those dominated by very large river systems (e.g., much of South America). Regions with recent tectonic activity or smaller, more isolated freshwater systems often are divided into smaller ecoregions. For example, central Mexico has experienced intermittent isolation and exchange between basins owing to active mountain-building processes leading to small, fragmented systems with distinct faunas. We acknowledge that data quality may also influence the size of ecoregions; for instance, the entire Amazon is currently divided into only 13 ecoregions, but better data on species occurrences within major subbasins would most likely support finer delineations.

| Region               | Delineation approach  |
|----------------------|---|
| Africa               | Using Roberts (1975) as a starting point, ecoregions were delineated using a top-down qualitative assessment that<br>incorporated expert knowledge and divisions of major river basins. In a few cases where basin divides do not circumscribe<br>species distributions or where basins contain internal barriers to dispersal, ecoregions straddle or divide basins.   |
| Middle East          | Species lists were generated for whole drainage basins, which were then either combined with smaller catchments that were very similar faunistically (minor desert basins, for example) or subdivided on the basis of different ecologies (e.g., th Tigris-Euphrates with lowland marshes and upland streams).  |
| Former USSR          | A species/genera/family presence/absence matrix was compiled for a hierarchy of hydrographic units, and cluster analysi<br>and ordination techniques (Primer v.6 statistics software) were employed to assess biotic similarities among hydrographic<br>units and to identify major faunal breaks.  |
| Remainder of Eurasia | For Southeast Asia and southern Europe, a bottom-up approach employing both published and unpublished field data and<br>expert assessment was used. East Asian, northern European, and eastern European ecoregions were delineated through<br>top-down process using major basins as a starting point and incorporating traditionally recognized zoogeographic patterns<br>where appropriate.   |
| Australasia          | For Australia, ecoregions were adapted from Allen and colleagues' (2002) and Unmack's (1991) "freshwater fish biogeo-<br>graphic provinces"; provinces were derived through similarity analyses, parsimony analysis, and drainage-based plots of<br>species ranges. For New Guinea, "subprovinces" of Allen (1991) were modified (primarily combined) on the basis of exper<br>input. For New Zealand and other islands and island groups, islands were placed in ecoregions on the basis of expert<br>input. |
| Oceania              | Islands and island groups were placed in ecoregions on the basis of distinctive (endemic or near-endemic) fish faunas.  |
| Canada               | Separate cluster analyses were conducted on fish occurrence in the secondary watersheds in each of the nine primary watersheds in Canada.   |
| United States        | The "subregions" of Maxwell and colleagues (1995) were adopted, with relatively small modifications made following inpu<br>by regional specialists, especially the Endangered Species Committee of the American Fisheries Society.  |
| Mexico               | Ecoregion delineations were based on qualitative similarity/dissimilarity assessments of major basins, using the standard<br>administrative hydrographical regions of the Mexican federal government. Subregions within major basins were recognized<br>as separate ecoregions when the fish fauna was sufficiently distinctive.  |
| Central America      | Fish provinces from Bussing (1976) were revised and subdivided on the basis of the application of the similarity index to subbasin fish presence/absence data.  |
| Caribbean            | Ecoregions from Olson and colleagues (1998) were modified on the basis of similarity analyses of island-by-island species lists and expert input.   |
| South America        | Ecoregion delineations were based on qualitative similarity/dissimilarity assessments of catchments, resulting in aggrega-<br>tion/disaggregation. See box 1 for additional information.  |

*Note:* Some of the variations resulting from differences in data quality and biogeographic drivers across and within regions are noted. For some regions, subecoregions (described at *www.feow.org*) were delineated to capture finer-scale patterns than could be represented by ecoregions.

## Box 1. Example of criteria applied to ecoregion delineation: South America.

The delineation process for South America followed a stepwise process of subdivision of the continent's major drainage systems. Delineation started with the historically recognized major ichthyographic provinces exemplified in Gery (1969) and Ringuelet (1975) and proceeded with subdivision at finer scales using regionalized data on fish distributions.

The criteria for determining the merit of delineating an ecoregion were not uniform across the continent as a result of localized faunistic differences. In some areas, delineations were based on family-level data, whereas in others, faunistic turnover at lower taxonomic levels was the criterion. For instance, astroblepid catfishes are distinct components of highelevation freshwaters along the Andes forefront, and that family's distribution was critical to informing the delineation of the high Andean ecoregions. On the other side of the continent along the Atlantic coast, we used the presence or absence of endemic assemblages of the genus *Trichomycterus*, several genera of the subfamily Neoplecostomatinae, and the presence or absence of annual killifish genera and species to distinguish distinct drainage complexes from one another.

In the piedmont zones and in contact areas between lowlands and geologic shield areas, we used indicator groups to determine where along the elevation/slope gradient the fauna was changing. The distribution of lowland forms was matched with forms found in higher-gradient systems to establish where one group was dropping out and the other started occurring. This transition zone was then established as the operational boundary between connecting ecoregions.

For areas like Patagonia, the Titicaca altiplano, and the Maracaibo basin, the uniqueness of the fauna, often occurring within clearly defined geographic areas, permitted reasonably straightforward delineations. In the larger river basin systems where there are no clear boundaries, the ecoregional limits are the best approximation, given the current data.

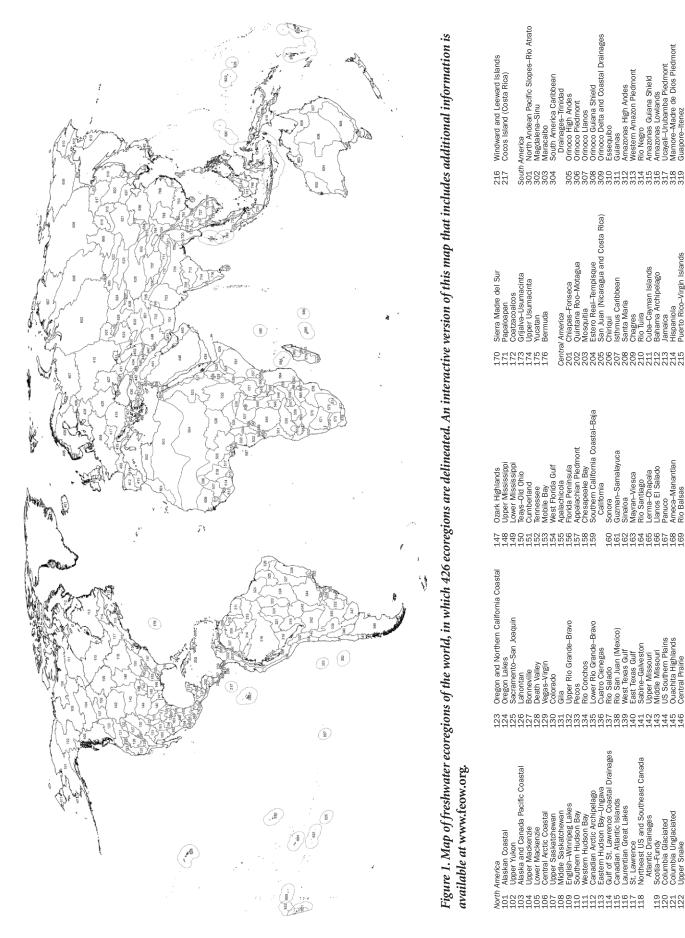
The process of delineating ecoregions required compiling and synthesizing information on the distributions of fish species. A logical and practical extension of the delineations was the compilation of fish species lists for each ecoregion. For the United States, NatureServe provided presence/absence data for individual species, coded to eight-digit hydrologic unit codes (HUCs); these HUC occurrences were then translated to ecoregions, and the data were manually cleaned of erroneous occurrences derived from introductions and problematic records. These species lists were then merged with those from Canada and Mexico for transnational ecoregions. For all other ecoregions, data came from the published literature, as well as from gray literature and unpublished sources (see table 1; a full bibliography is available at *www.feow.org*). In all cases, experts served as gatekeepers of these data to ensure that lists were based on the best available information, both in terms of distributions and nomenclature. Introduced species were removed from the tallies presented here, as were undescribed species. Confirmed extinct species (Ian J. Harrison, American Museum of Natural History, New York, personal communication, 29 March 2007) were excluded, but extirpated species were included to acknowledge restoration opportunities. Endemic species, defined as those occurring only in a single ecoregion, were identified first by experts and cross-checked using a species database constructed for this project, which includes more than 14,500 described fish species. Species were coded as freshwater, brackish, or marine using data from FishBase (*www.fishbase.org*), and species with only brackish or marine designations were omitted from the richness and endemism totals reported here.

#### Freshwater ecoregional map and species results

Our map of freshwater ecoregions contains 426 units, covering nearly all nonmarine parts of the globe, exclusive of Antarctica, Greenland, and some small islands (figure 1; a full legend is available at *www.feow.org*). There is large variation in the area of individual ecoregions. Large ecoregions, such as the dry Sahel (4,539,429 km<sup>2</sup>), tend to be found in more depauperate desert and polar regions exhibiting low species turnover. Smaller ecoregions are typically found in noncontinental settings where systems are by nature smaller and species turnover is higher, as in the Indo-Malay region. The smallest ecoregion, at 23 km<sup>2</sup>, is Cocos Island (Costa Rica); the average ecoregion size is 311,605 km<sup>2</sup>. Ecoregions ranged from those encompassing only 1 country to those straddling 16 countries (central and western Europe ecoregion).

In total, we assigned more than 13,400 described freshwater fish species to ecoregions, of which more than 6900 were assigned to single ecoregions (i.e., endemic). Examination of the fish species data synthesized by ecoregion confirms some well-known patterns and highlights others unknown to many conservationists, managers, and policymakers working at regional or global scales (figures 2a-2d). In agreement with previous global assessments (Groombridge and Jenkins 1998, Revenga et al. 1998), our analysis identifies as outstanding for both fish richness and endemism systems that include large portions of Africa's Congo basin, the southern Gulf of Guinea drainages, and Lakes Malawi, Tanganyika, and Victoria; Asia's Zhu Jiang (Pearl River) basin and neighboring systems; and large portions of South America's Amazon and Orinoco basins. Areas confirmed for globally high richness include Asia's Brahmaputra, Ganges, and Yangtze basins, as well as large portions of the Mekong, Chao Phraya, and Sitang and Irrawaddy; Africa's lower Guinea; and South America's Paraná and Orinoco. When richness is adjusted for ecoregion area, additional systems such as the Tennessee, Cumberland, Mobile Bay, Apalachicola, and Ozark highlands in the southeastern United States; portions of Africa's Niger River Basin; the islands of New Caledonia, Vanuatu, and Fiji; China's Hainan Island; and large parts of Sumatra and Borneo, among many other areas, are also especially noteworthy.

Numerous systems previously identified as highly endemic for fish were confirmed, as measured by either numbers of endemic species or percentage endemism. A subset includes



Bight Drainages Northern Guir of Guinea Drainages-Bioko Nestern Eduatorial Crater Lakes Lake Crad Lake Victoria Basin Upper Nile Lower Nile Nile Delta Turan Plain Northem Hormuz Drainages Caspian Mairtem Caspian Drainages Volga Delta-Northern Caspian Drainages Eastern Zimbabwe Highlands Coastal East Affica Lake Rukwa Southern Eastern Rift Tana, Athi, and Coastal Drainages Dgooue-Nyanga-Kouilou-Niari Southern Gulf of Guinea Drainages and Madagascar Atlantic Northwest Africa Mediterranean Northwest Africa Karoo Drakensberg-Maloti Highlands Vestern Red Sea Drainages Vorthern Eastern Rift Zambezian Headwaters Jpper Zambezi Floodplains Sahara Diy Saha Lower Niger-Benue Niger Den Niger Upper Niger Energa Cambia Fouta-Djalon Northern Upper Guinea Southern Upper Guinea Aiddle Zambezi-Luangwa Sudanic Congo-Oubangi Upper Congo Rapids Upper Congo Albertine Highlands Lake Tanganyika Malagarasi-Moyowosi Bangweulu-Mweru Upper Lualaba Aavir and Lut Deserts Malebo Pool Lower Congo Rapids Lower Congo Arstveld Sink Holes Zambezian Highveld ower Zambezi Ethiopian Highlands ...ake Tana Southern Kalahari Western Orange **Duvette Centrale** Horn of Africa Lake Turkana Shebelle-Juba Ndombe ake Malawi /ango Eburneo Ashanti Aulanje **Kalahar**i Vamak Sangha Suanza Etosha imba Vamib Mai 447 450 451 451 451 451 452 453 Northeastern Caatinga and Coastal Upper Uruğuay Laguna dos Patos Tramandai-Mampituba Central Andean Pacific Slopes S. Francisco Northeastern Mata Atlantica Paraiba do Sul Ribeira de Iguape Southeastern Mata Atlantica azonas Estuary and Coastal Western Transcaucasia Kura-South Caspian Drainages Southwestern Arabian Coast Arabian Interior Lower Tigris and Euphrates Upper Tigris and Euphrates Atacama Mar Chiquita-Salinas Grandes Caritabric Coast-Languedoc Centrabric Coast-Languedoc Central and Western Europe Norwegian Sea Drainages Northern Baltic Drainages Barents Sea Drainages Southern Baltic Lowlands Lake Onega-Uai Volga-Uai Western Caspian Drainages Cuyan-Desaguadero South Andean Pacific Slopes Gulf of Venice Drainages Italian Peninsula and Islands Dalmatia Southeast Adriatic Drainages Bonder Bonder Patagonia Valdivian Lakes Guapagos Islands Guan Fernandez Island Fuminense Tapajos-Juruena Madeira Brazilian Shield Europe and Middle East 402. Iceland-Jam Mayen 403. Iceland-Jam Mayen 403. Northern British Isles 405. Northern British Isles 405. Norwegian Sea Drainage 405. Norwegian Sea Drainage 407. Barents Sea Drainage 409. Lake Ornega-Jank Ladog 410. Voiga-Jural 411. Western Iberia 413. Southen Breria 414. Eastern Iberia 415. Duper Danube 416. District Arainages 416. District Arainages 416. District Arainages 417. Upper Danube 418. Dinester-Lower Danube 419. Dinester-Lower Danube 420. Southensula and Isla 411. Upper Danube 422. Vardar 423. Northern Aratolia 424. Mestern Anatolia 423. Western Anatolia 423. Western Anatolia 423. Western Anatolia 423. Western Anatolia 423. Southwestern Anatolia 423. Southwestern Anatolia 423. Western Anatolia 423. Southwestern Anatolia 423. Southwestern Anatolia 423. Southwestern Anatolia 423. Southwestern Anatolia 424. Oromyka 425. Connea River 423. Southwestern Anatolia 424. Oromyka 425. Conta River 426. Conta River 427. Doner 428. Southwestern Anatolia 428. Jordan River 429. Oromyka 420. Oromika 420. Oromika 420. Oromika 420. Oromika 420. Oromika Dniester-Lower Danube Drainages Tocantins-Araguaia Paraguay Upper Parana Lower Parana -ower Uruguay Drainages Parnaiba guassu Titicaca C) Paco AB 321 321 323 323 325 325 326 

Upper Indus Tibetan Plateau Endorheic Drainages Western Madagascar Wortwestern Madagascar Madagascar Eastern Highlands Southern Madagascar Madagascar Eastern Lowlands Comoros-Mayotte Songhua Jiang Inner Mongolia Endomeic Basins Western Mongolia Lower and Middle Syr Darya Lake Issyk Kul-Upper Chu Northern Central Asian Highlands Hamgyong-Sanmaek Sakhalin, Hokkaido, and Sikhote-Kamchatka and Northern Kurils Okhotsk Coast Coastal Amur Eastern Yellow Sea Drainages Southeastern Korean Peninsula Mascarenes S. Tome and Principe–Annobon Zambezian Lowveld Amatolo-Winterberg Highlands Namuda-Tapi Ganges Detta and Plain Ganges Himalayan Fodhills Upper Brahmaputra Middle Brahmaputra Northern Deccan Plateau Southern Deccan Plateau Western Ghats. Southern Temperate Highveld Southeastern Ghats Southeastern Ghats Sri Lanka Wet Zone Sri Lanka Wet Zone Chin Hilka-Arakan Coast Upper Salween Lower and Middle Salween Upper Huang He Upper Huang He Corridor Huang He Great Bend Lower Huang He Yaghistan Indus Himalayan Foothills Alin Coast Honshu-Shikoku-Kyushu Biwa Ko Baluchistan Helmand–Sistan Lower and Middle Indus Aral Sea Drainages Middle Amu Darya Upper Amu Darya Sea Drainages Dzungaria Balkash–Alakul Anadyr East Chukotka Lower Amur Middle Amur Irgyz-Turgai Ob Shilka (Amur) Yenisei Lake Baikal Jpper Irtysh Seychelles Cape Fold Kolyma Korvakia -iao He Southern Asia Southern Asia 702 Beluchis Helmann 703 Lower Helmann 707 Helmann 707 Helmann 706 Unper In 707 Hoten 706 Unper ID 708 Namuda 708 Namuda 709 Namuda 709 Namuda 700 Unper ID 711 Unper ID 712 Nuther 713 Souther 713 Souther 713 Souther 713 Souther 715 Souther 717 Souther 718 Souther 717 Souther 718 Souther 71 Chuya aimvr Qaidan ern Asia arim 8440 642 643

Vogelkop-Bomberai New Guinea North Coast New Guinea Central Mountains Southwest New Guinea-Trans-Fly Lowland Papuan Peninsula Bismarck Antineula Bismarck Antineulago Solomon Islands Southern Cean Slope of Sumatra and Java Southern Central Sumatra Southern Sumatra-Western Java Central and Eastern Java Mae Khlong Malay Peninsula Eastern Slope Northern Central Sumatra-Western Southern Annam Eastern Gulf of Thailand Drainages Chao Phraya Lower Lancang (Mekong) Khorat Plateau (Mekong) Kratie–Stung Treng (Mekong) Mekong Delta Lake Lanao Northern Philippine Islands Palawan-Busuanga-Mindoro Western Taiwan Eastern Taiwan Inle Lake Upper Lancang (Mekong) Er Hai Arafura-Carpentaria Lake Eyre Basin Eastern Coastal Australia Lower Yang Ze Coastal Fujian-Zeijang Andaman Islands Nicobar Islands Southwestern Australia Pilbara Borneo Highlands Northeastern Borneo Eastern Borneo Southeastern Borneo Hawaiian Islands East Caroline Islands West Caroline Islands Malukku Lesser Sunda Islands Murray–Darling Bass Strait Drainages Northwestern Borneo Society Islands Tubuai Islands Marquesas Islands Southern Tasmania Northern Annam Xi Yiang Upper Yangtze Middle Yangtze Vanuatu New Caledonia Song Hong Yunnan Lakes Sulawesi Malili Lakes Lake Poso Mallis-Futuna and Pacific New Zealand Malaysia Kimberley Mindanao Hainan Samoas (apuas Paleo Rapa Acet 8.821 Section 2.2 Section 2.2

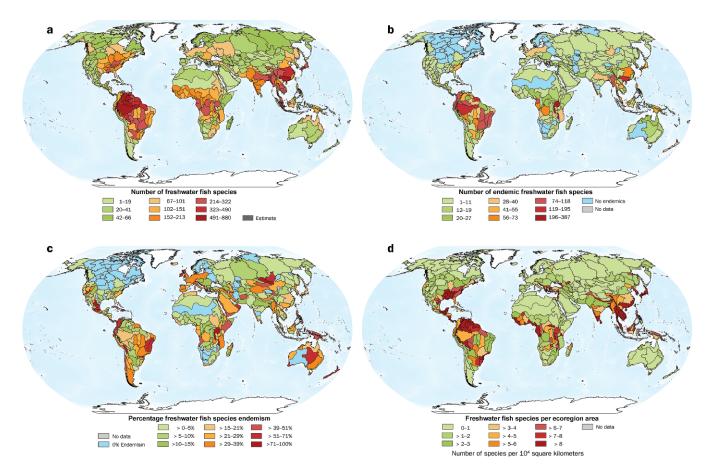


Figure 2. Preliminary freshwater fish species data for ecoregions: (a) species richness, (b) number of endemic species, (c) percentage endemism, and (d) species per ecoregion area. Numbers may be adjusted on the basis of an ongoing process to correct nomenclatural errors. Natural breaks (Jenk's optimization) was the classification method used for panels (a)–(c). This method identifies breakpoints between classes using a statistical formula that identifies groupings and patterns inherent in the data.

highland lakes in Cameroon along with Africa's Lake Tana; northwestern and eastern Madagascar; freshwaters from Turkey's central Anatolia region, the northern British Isles, the Philippines, Sri Lanka, India's western Ghats, the southwestern Balkans, and northwest Mediterranean; southwestern Australia and nearly the entire island of New Guinea; Eurasian lakes, including Baikal, Inle, and Sulawesi's Lake Poso and Malili system; Death Valley in the United States and Mexico's Pánuco system; and South America's Iguaçu River, Lake Titicaca, and the freshwaters of both the Mata Atlántica and the continent's northwestern Pacific coast. Additionally, newly available data show that some systems previously recognized for high endemism, such as those of South America's Guianas, also exhibit exceptional richness.

Because our ecoregions cover all nonmarine waters, and because they often exist as subdivisions of major river basins, our results also highlight a number of smaller systems for the first time in global analyses. Using finer-resolution data allowed us to identify the high richness of the Congo's Malebo Pool and Kasai basin. Cuba and Hispaniola stand out for endemism, along with the Amazon's western piedmont and the Tocantins-Araguaia systems. The Tocantins-Araguaia, as well as the highly endemic São Francisco, were defined as units of analysis in Revenga and colleagues (1998), but fish data were unavailable for those basins when that study was done. Systems never before analyzed globally but recognized in our results as exceptionally rich for fish include those of the Malay Peninsula's eastern slope and Japan. A large number of ecoregions are identified for the first time for highly endemic faunas, measured as percentage endemism. Newly identified ecoregions with at least 50% endemism include Africa's Cuanza, Australia's Lake Eyre Basin, Mexico's Mayrán-Viesca, and New Zealand, as well as a large number of highly depauperate ecoregions such as Africa's karstveld sink holes, Turkey's Lake Van, the Oman Mountains, western Mongolia, and Hawaii.

Each of the biodiversity analyses that we offer here emphasizes different sets of ecoregions, suggesting that a single measure of species diversity might overlook ecoregions of important biodiversity value. In a comparative analysis of biodiversity value, ecoregions are probably best evaluated against others within the same region, with similar historical and environmental characteristics, and of similar size to account for the typically positive relationship between river discharge and fish species richness (Oberdorff et al. 1995). Nonetheless, some systems, such as the Amazon and many of Africa's Rift Valley lakes, stand out by nearly any measure of fish biodiversity and are indisputable global conservation priorities.

#### **Conservation applications**

The ecoregion map and associated species data summarized here have a number of conservation applications. At global and regional scales the ecoregion map can be used to distinguish distinct units of freshwater biodiversity to be represented in conservation efforts. The Convention on Wetlands, for instance, requires that sites nominated as wetlands of international importance-with wetlands defined to include all freshwaters-be evaluated against a "biogeographic regionalization" criterion (Ramsar Bureau 2006). Lack of a global biogeographic scheme has stalled the application of this criterion, but our global map and database may provide a necessary framework for identifying broadscale gaps in protection. Similarly, progress toward the establishment of representative networks of freshwater protected areas, as called for by the third IUCN World Conservation Congress, the fifth World Parks Congress, and the seventh Meeting of the Conference of the Parties to the Convention on Biological Diversity, can now be measured using ecoregions as a proxy for finer-scale global species or habitat distribution data. At a regional level, the freshwater ecoregion map may be used as supplementary information for implementation of the European Union's Water Framework Directive (2000/60/EC), which requires a characterization of surface water bodies and currently uses regions defined by Illies (1978).

A primary use of ecoregions is as conservation planning units (Higgins 2003). Our attribution of freshwater fish species data to ecoregions is an important first step for datapoor regions. Organizations or agencies with regional mandates may choose to compare biodiversity values across ecoregions in the process of setting continental priorities (Abell et al. 2000, Thieme et al. 2005). At the basin scale, ecoregions can help to introduce biodiversity information into water-resource or integrated-basin management activities (Gilman et al. 2004). Where major basins are divided among multiple freshwater ecoregions, whole-basin exercises can use ecoregions as stratification units to ensure adequate representation of distinct biotas. Where unconnected drainages are combined into a single freshwater ecoregion, planners may choose to consider a counterintuitive planning unit to incorporate biogeographic patterns. Freshwater ecoregions defined in previous exercises have already been put to use by the Nature Conservancy and WWF in numerous conservation planning efforts across North America (e.g., Upper Mississippi; Weitzell et al. 2003), South America (e.g., the Pantanal; de Jesus 2003), and Africa (e.g., the Congo basin; Kamdem-Toham et al. 2003).

#### **Caveats and limitations**

Ecoregions are delineated based on the best available information, but data describing freshwater species and ecological processes are characterized by marked gaps and variation in quality and consistency. Data quality is generally considered high for North America, Australia, New Zealand, Japan, western Europe, and Russia; moderate for Central America, the southern cone of South America, southern and western Africa, Oceania, and the Middle East; and poor for much of southeastern Asia, central and eastern Africa, and South America north of the Paraná River basin.

Freshwater ecoregions are not homogeneous units. Within individual ecoregions there will be turnover of species along longitudinal gradients of river systems and across different habitats such as flowing and standing-water systems. The inclusion of multiple macrohabitat types within a given freshwater ecoregion is a marked departure from terrestrial ecoregions, which typically encompass a single vegetation-defined biome (e.g., deciduous forests, evergreen forests, or scrub; Wikramanayake et al. 2002).

Ecoregions are imperfect units for highlighting certain highly distinct and highly localized assemblages occurring at subecoregion scales. Examples include many peat swamps or subterranean systems. Underground systems such as caves and karsts may require their own planning framework, as groundwater catchments may not correspond with the surface-water catchments upon which our ecoregions are built.

For reasons of practicality and scale, our ecoregion framework does not take into account the distributions of freshwater species such as invertebrates, reptiles, and amphibians. This is a limitation of the ecoregional approach presented here, which is especially problematic for places such as isolated islands where freshwater fish provide little information to inform biogeographic delineation. We hope this taxonomic omission will serve as motivation for generating and synthesizing global data for other taxonomic groups to provide complementary information for conservation planners, particularly when working at subecoregional scales. We recognize that improved information in the future may warrant map revisions, and we highlight areas of greatest data uncertainty in part to encourage enhanced research investment in those places. We believe that the critical state of freshwater systems and species argues against waiting for ideal biodiversity data to be developed before generating urgently needed conservation tools like the ecoregion map.

Shifting transition zones for species are common, and we recommend that ecoregions be viewed as logical units for more detailed analyses and strategies. Ecoregions are intended to depict the estimated original extent of natural communities before major alterations caused by recent human activities, but original distributions can be difficult to reconstruct. As new species are described, our understanding of distribution patterns may also change. Ecoregional delineation is an iterative process, and changes to ecoregion boundaries should be incorporated as new information becomes available. There is no definitive, error-free data source for classifying fish species as freshwater, brackish, or marine. We chose to use the global FishBase habitat assignments, which are derived from the literature, to ensure that any given species in our database would be classified consistently wherever it occurred. We recognize that errors of omission or commission may derive from inaccuracies in the FishBase assignments as well as from the habitat plasticity of some species. All species information provided to us by experts, regardless of habitat assignment, is retained in our database for future analyses.

The preliminary richness and endemism numbers presented here are in some cases markedly different from existing estimates in the literature. For example, our tally for Lake Malawi contains 431 described fish species, but other estimates run as high as 800 or more (Thieme et al. 2005). Our omission of undescribed species, as well as the conservative approach taken by experts in using only robust species occurrence data, account for many of these lower-thanexpected numbers. Numbers of endemics may in some cases be higher than expected because endemics were identified strictly through a database query for unique occurrences, and many species lists are undoubtedly incomplete or use synonyms. We anticipate that many tallies will change with further refinement of species lists but that the broad patterns presented here will hold.

#### **Conclusions**

The newly available species data attributed to ecoregions has important implications for prioritizing conservation investments. As one illustration, in 2005 the Global Environment Facility (GEF), which spends more than \$1 billion each year on environmental projects, adopted a new resource allocation framework. Terrestrial ecoregion maps and biodiversity data were notable inputs to the framework, but parallel freshwater information to help guide investments was lacking. The GEF framework fortunately leaves open the possibility of incorporating freshwater ecoregions and biodiversity data at a later date (GEF 2005).

In addition to providing data for scientific and conservation purposes, we aim to give the largest possible number of people access to the ecoregion-level information collected in association with the global map. The information will be freely available on the Internet (*www.feow.org*) as well as in brochures, posters, and other publications. The freshwater ecoregion map covers virtually all land surfaces on Earth, so people around the globe will have the opportunity to learn about the freshwater systems where they live.

For most policymakers, water resource managers, and even conservationists, freshwater biodiversity is more of an afterthought than a central consideration of their work. The freshwater ecosystem services that support the lives and livelihoods of countless people worldwide are a far larger concern. Yet freshwater biodiversity and ecosystem services are linked through ecological integrity, and better-informed efforts to conserve freshwater biodiversity should benefit human communities as well. The freshwater ecoregions of the world map and associated species data begin to improve access to previously dispersed and difficult to access freshwater biodiversity information. We hope that this set of products catalyzes additional work toward a better understanding of freshwater species distributions and—of equal if not more importance leads to a ramping up of freshwater conservation activity and success.

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#### **References cited**

- Abell RA, et al. 2000. Freshwater Ecoregions of North America: A Conservation Assessment. Washington (DC): Island Press.
- Allen GR. 1991. Field Guide to the Freshwater Fishes of New Guinea. Madang (Papua New Guinea): Christensen Research Institute.
- Allen GR, Midgley SH, Allen M. 2002. Field Guide to the Freshwater Fishes of Australia. Perth (Australia): Western Australian Museum.
- Balian EV, Lévêque C, Segers H, Martens K, eds. 2008. Freshwater Animal Diversity Assessment. Dordrecht (Netherlands): Springer.
- Banarescu P. 1990. Zoogeography of Fresh Waters, vol. 1: General Distribution and Dispersal of Freshwater Animals. Weisbaden (Germany): AULA.
- Berra TM. 2001. Freshwater Fish Distribution. San Diego: Academic Press. Burgess GH, Franz R. 1989. Zoogeography of the Antillean freshwater fish
- fauna. Pages 263–304 in Woods CA, Sergile FE, eds. Biogeography of the West Indies: Patterns and Perspectives. Boca Raton (FL): CRC.
- Bussing WA. 1976. Geographic distribution of the San Juan ichthyofauna of Central America with remarks on its origin and ecology. Pages 157–175 in Thorson TB, ed. Investigations of Nicaraguan Lakes. Lincoln: University of Nebraska.
- Contreras-Balderas S. 2000. Biogeografía mexicana de peces continentales. Mexicoa 2: 80–84.
- de Jesus F, coord. 2003. Classification of Aquatic Ecosystems of the Pantanal and the Upper Paraguay Watershed. Brasilia (Brazil): Nature Conservancy.
- Dinerstein E, Olson DM, Graham DJ, Webster AL, Primm SA, Bookbinder MP, Ledec G. 1995. A Conservation Assessment of the Terrestrial Ecoregions of Latin America and the Carribean. Washington (DC): World Bank.
- Dudgeon D, et al. 2006. Freshwater biodiversity: Importance, threats, status and conservation challenges. Biological Reviews 81: 163–182.
- Eschmeyer WN. 2006. The Catalog of Fishes On-line (updated 7 November 2006). California Academy of Sciences. (14 March 2008; *www.calacademy. org/research/ichthyology/catalog/fishcatsearch.html*)
- [GEF] Global Environment Facility. 2005. Technical Paper on the GEF Resource Allocation Framework. (4 April 2008; www.gefweb.org/ Operational\_Policies/Resource\_Allocation\_Framework.html)

- Gery J. 1969. The fresh-water fishes of South America. Pages 828–848 in Fitkau EJ, ed. Biogeography and Ecology in South America. The Hague (Netherlands): W. Junk.
- Gilman RT, Abell RA, Williams CE. 2004. How can conservation biology inform the practice of integrated river basin management? Journal of River Basin Management 2: 135–148.
- Groombridge B, Jenkins M. 1998. Freshwater Biodiversity: A Preliminary Global Assessment. Cambridge (UK): World Conservation Monitoring Centre.
- Groves CR, Jensen DB, Valutis LL, Redford KH, Shaffer ML, Scott JM, Baumgartner JV, Higgins JV, Beck MW, Anderson MG. 2002. Planning for biodiversity conservation: Putting conservation science into practice. BioScience 52: 499–512.
- Higgins JV. 2003. Maintaining the ebbs and flows of the landscape: Conservation planning for freshwater ecosystems. Pages 291–318 in Groves C, ed. Drafting a Conservation Blueprint: A Practitioner's Guide to Planning for Biodiversity. Washington (DC): Nature Conservancy and Island Press.
- Illies J, ed. 1978. Limnofauna Europaea. New York: Gustav Fischer.
- Kamdem-Toham A, et al. 2003. Biological Priorities for Conservation in the Guinean-Congolian Forest and Freshwater Region. Libreville (Gabon): WWF-CARPO.
- Keith P, Vigneux E, Marquet G. 2002. Atlas des Poissons et des Crustacés d'Eau Douce de Polynésie Française. Paris: Muséum National d'Histoire Naturelle.
- Kottelat M, Freyhof J. 2007. Handbook of European Freshwater Fishes. Switzerland: Steven Simpson Books.
- Lévêque C, ed. 1997. Biodiversity Dynamics and Conservation: The Freshwater Fish of Tropical Africa. Cambridge (UK): Cambridge University Press.
- Lévêque C, Balian EV, Martens K. 2005. An assessment of animal species diversity in continental waters. Hydrobiologia 542: 39–67.
- Lundberg JG, Kottelat M, Smith GR, Stiassny MLJ, Gill AC. 2000. So many fishes, so little time: An overview of recent ichthyological discovery in continental waters. Annals of the Missouri Botanical Garden 87: 26–62.
- Maxwell JR, Edwards CJ, Jensen ME, Paustian SJ, Parrott H, Hill DM. 1995. A Hierarchical Framework of Aquatic Ecological Units in North America (Nearctic Zone). St. Paul (MN): USDA Forest Service, North Central Forest Experiment Station. General Technical Report NC-176.
- McDonald R, McKnight M, Weiss D, Selig E, O'Connor M, Violin C, Moody A. 2005. Species compositional similarity and ecoregions: Do ecoregion boundaries represent zones of high species turnover? Biological Conservation 126: 24–40.
- McDowall RM. 1990. New Zealand Freshwater Fishes: A Natural History and Guide. Auckland (New Zealand): Hinemann Reed.
- [MEA] Millennium Ecosystem Assessment. 2005. Ecosystems and Human Well-being: Synthesis. Washington (DC): Island Press.
- Menni RC. 2003. Peces y ambientes en la Argentina continental. Monograph of the Museo Argentino de Ciencias Naturales 5: 1–316.
- Miller RR, Minckley WL, Norris S. 2005. Fishes of México. Chicago: University of Chicago Press.
- NatureServe. 2006. NatureServe Explorer: An Online Encyclopedia of Life. Version 6.1. NatureServe. (17 March 2008; www.natureserve.org/explorer)
- Oberdorff T, Guegan JF, Hugueny B. 1995. Global scale patterns of fish species richness in rivers. Ecography 18: 345–352.
- Olson DM, Dinerstein E, Canevari P, Davidson I, Castro G, Morisset V, Abell R, Toledo E. 1998. Freshwater Biodiversity of Latin America and the Caribbean: A Conservation Assessment. Washington (DC): Biodiversity Support Program.
- Olson DM, et al. 2001. Terrestrial ecoregions of the world: A new map of life on Earth. BioScience 51: 933–938.
- Omernik JM. 1987. Ecoregions of the conterminous United States. Annals of the Association of American Geographers 77: 118–125.
- Paavola R, Muotka T, Virtanen R, Heino J, Jackson D, Maki-Petays A. 2006. Spatial scale affects community concordance among fishes, benthic macroinvertebrates, and bryophytes in streams. Ecological Applications 16: 368–379.

### Articles

- Ramsar Bureau. 2006. Strategic Framework for the List of Wetlands of International Importance of the Convention on Wetlands (Ramsar, Iran, 1971), edition 2006. (17 March 2008; *www.ramsar.org/key\_guide\_ list2006\_e.htm*)
- Rauchenberger M. 1988. Historical biogeography of Poeciliid fishes in the Caribbean. Systematic Zoology 37: 356–365.
- Reis R, Kullander S, Ferraris C, eds. 2003. Check List of the Freshwater Fishes of South and Central America. Porto Alegre (Brazil): EDIPUCRS.
- Revenga C, Kura Y. 2003. Status and trends of biodiversity of inland water ecosystems. Montreal (Canada): Secretary of the Convention on Biological Diversity. Technical Series no. 11.
- Revenga C, Murray S, Abramovitz J, Hammond A. 1998. Watersheds of the World: Ecological Value and Vulnerability. Washington (DC): World Resources Institute.
- Revenga C, Campbell I, Abell R, de Villiers P, Bryer M. 2005. Prospects for monitoring freshwater ecosystems towards the 2010 targets. Philosophical Transactions of the Royal Society B 360: 397–413.
- Ringuelet R. 1975. Zoogeografía y ecología de los peces de aguas continentales de la Argentina y consideraciones sobre las áreas ictiológicas de América del Sur. Ecosur 2: 1–122.
- Roberts TR. 1975. Geographical distribution of African freshwater fishes. Zoological Journal of the Linnean Society 57: 249–319.
- Scott WB, Crossman EJ. 1998. Freshwater Fishes of Canada. Oakville (Canada): Galt House.
- Skelton PH. 1994. Diversity and distribution of freshwater fishes in East and Southern Africa. Annals of the Royal Central Africa Museum (Zoology) 275: 95–131.
- Skelton PH, Cambray JA, Lombard A, Benn GA. 1995. Patterns of distribution and conservation status of freshwater fishes in South Africa. South African Journal of Zoology 30: 71–81.
- Spalding MD, et al. 2007. Marine ecoregions of the world: A bioregionalization of coast and shelf areas. BioScience 57: 573–583.

- Thieme ML, Abell R, Stiassny MLJ, Skelton P, Lehner B, Teugels GG, Dinerstein E, Kamdem-Toham A, Burgess N, Olson D. 2005. Freshwater Ecoregions of Africa and Madagascar: A Conservation Assessment. Washington (DC): Island Press.
- Tonn WM. 1990. Climate change and fish communities: A conceptual framework. Transactions of the American Fisheries Society 119: 337–352.
- Udvardy MDF. 1975. A Classification of the Biogeographic Provinces of the World. Gland (Switzerland): IUCN. Occasional Paper no. 18.
- Unmack PJ. 2001. Biogeography of Australian freshwater fishes. Journal of Biogeography 28: 1053–1089.
- Vinson MR, Hawkins CP. 2003. Broad-scale geographical patterns in local stream insect genera richness. Ecography 26: 751–767.
- Wallace AR. 1876. The Geographical Distribution of Animals. New York: Harper.
- Wasson J-G, Barrera S, Barrere B, Binet D, Collomb D, Gonzales I, Gourdin F, Guyot J-L, Rocabado G. 2002. Hydro-ecoregions of the Bolivian Amazon: A geographical framework for the functioning of river ecosystems. Pages 69–91 in McClain ME, ed. The Ecohydrology of South American Rivers and Wetlands. Wallingford (UK): International Association of Hydrological Sciences.
- Weitzell RE, Khoury ML, Gagnon P, Schreurs B, Grossman D, Higgins J. 2003. Conservation Priorities for Freshwater Biodiversity in the Upper Mississippi Basin. Arlington (VA): NatureServe and The Nature Conservancy.
- Wikramanayake E, Dinerstein E, Loucks C, Olson D, Morrison J, Lamoreux J, McKnight M, Hedao P. 2002. Ecoregions in ascendance: Reply to Jepson and Whittaker. Conservation Biology 16: 238–243.

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