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FISH SPECIES RICHNESS IN RELATION TO RESTORED OYSTER REEFS, PIANKATANK RIVER, VIRGINIA

Juliana M. Harding and Roger Mann

Fish assemblages in relation to "reef" structures in marine habitats have been and continue to be topics for research addressing ecological and management questions. Much effort has been spent describing and defining fish assemblages, or groups of species, associated with tropical coral reefs (e.g., Sale 1991 and chapters therein), temperate hard bottom or rocky reefs (e.g., Sedberry and Van Dolah, 1984; Ambrose and Swarbrick, 1989), tropical lava flows (e.g., Godwin and Kosaki, 1989), and artificial "fishing" reefs (e.g., Chandler et al., 1985; Hueckel and Buckley, 1987; Bohnsack, 1989; Feigenbaum et al., 1989; Rountree, 1989; Stephan and Lindquist, 1989). Temperate oyster reefs, another natural reef type, host diverse finfish assemblages that are just beginning to be described (e.g., Wenner et al. 1996; Mann and Harding, 1997; Luckenbach et al. 1998 and references therein).

Before Eastern oyster (*Crassostrea virginica*) populations were significantly reduced by environmental degradation, fishing pressure, and disease, oyster reefs dominated the intertidal areas of Chesapeake Bay and supported complex ecological communities including many fish species. The living shell matrix created by these predominantly intertidal, estuarine reefs provides structural heterogeneity and vertical relief that attract and sustain fishes from many trophic levels similar to living coral reefs (Roberts and Ormond, 1987; Ebeling and Hixon, 1991; Friedlander and Parrish, 1998). Recreationally and commercially valuable piscivorous finfishes including striped bass (*Morone saxatilis*), bluefish (*Pomatomus saltatrix*), and weakfish (*Cynoscion regalis*) were and are integral components of trophic networks that depend on oyster reefs (Mann and Harding, 1997, 1998). These pelagic finfishes use oyster reefs as both feeding and nursery grounds (e.g., Breitburg, 1998; Mann and Harding, 1997, 1998; J. Harding and R. Mann, unpubl. data).

Current restoration efforts for both oysters and finfishes are realizing the interconnections between these species in the context of intertidal estuarine habitats (Wenner et al. 1996; Breitburg, 1998; Coen et al., 1997; Mann and Harding, 1997, 1998; Nestlerode et al., in preparation). Description of oyster reef fish assemblages and species richness over time provides a direct method to assess the importance of oyster reefs as fish habitat. As reef communities develop at restoration sites, the associated ichthyofauna should increase in species richness, i.e., the number of different species observed at a site. The objectives of this study were to qualitatively describe the ichthyofauna associated with oyster reef restoration sites in the Piankatank River, Virginia along a seasonal gradient and to compare observed fish assemblages and species richness in relation to species richness patterns from similar habitats.

METHODS AND RESULTS

STUDY SITE.—Field work was conducted at Palace Bar Reef, Piankatank River, Virginia (37°31'41.69"N, 76° 22'25.98"W) during 1996 and 1997 and was expanded to include

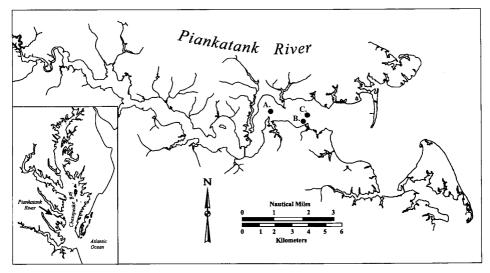


Figure 1. Map of the Piankatank River in relation to the Chesapeake Bay showing sampling locations. Palace Bar Reef (C.) was sampled in 1996 and 1997. Sites at Ginney Point (A.) and Roane Point (B.) were established in 1997 to provide data for non-reef habitats.

two non-reef sites within the Piankatank River during 1997 (Fig. 1). Palace Bar Reef is an intertidal oyster reef (210×30 m, reef depth range of 0.5 m above MLW to 3 m below MLW) that was built in July, 1993 adjacent to the historic Palace Bar oyster grounds. Approximately 70% of the reef (0.63 ha) is composed of oyster shell, while the remaining area (0.27 ha) is crushed clam shell (see Bartol and Mann, 1997 for a detailed site description). Since its construction in 1993, Palace Bar Reef has received annual oyster spat settlement (Bartol and Mann, 1997; J. Wesson, unpubl. data) and all oysters on the reef are due to natural settlement and recruitment, i.e., the reef was not initially seeded with oysters. The reef's oyster population is surveyed annually: oysters from multiple (n > 12 annually) 0.25 m quadrats placed randomly on the reef surface are collected by divers and subsequently counted and measured. Annual dredge surveys of the immediately adjacent Palace Bar are part of the VIMS Molluscan Ecology oyster monitoring program. These dredge surveys provide abundance and size class information for a natural oyster population immediately adjacent to Palace Bar Reef (R. Mann, unpubl. data). The first observation of adult or "market" oysters on the reef was during the fall of 1995.

In 1996, a trophic monitoring program was initiated for the reef area to provide baseline upper level trophic data (this study; Mann and Harding, 1997; Mann and Harding, 1998; Harding and Mann, in review). To evaluate reef versus non-reef habitat use by upper level consumers two non-reef sites within the Piankatank River were included in 1997. Data were collected at all sites during the same seasonal, temporal, and tidal scales. The non-reef site at Ginney Point (37° 31′ 52.78″N, 76°24′08.40″W) is located on a natural oyster shell bar (approximately 400×50 m, depth range 2 to 5 m; Fig. 1). The second non-reef site at Roane Point (37°31′37.48″N, 76°22′39.63″W) includes a sand bar south and inshore of Palace Bar Reef (approximately 400×15 m; depth range 2–4 m; Fig. 1). Mean tidal range in the Piankatank River is approximately 0.4 m. Water temperature and salinity were recorded once a week in conjunction with other monitoring from May to October during 1996 and 1997 at Ginney Point and Palace Bar Reef (Fig. 2). Water samples were taken

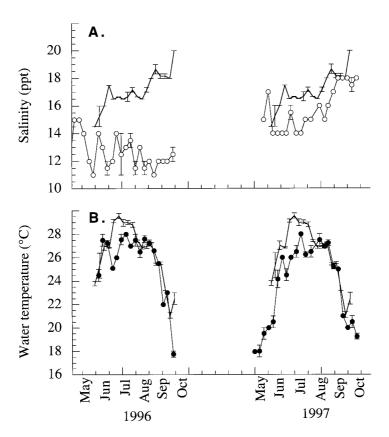


Figure 2. Mean salinity (‰, A.) and water temperature (°C, B.) values (\pm SE) for Ginney Point and Palace Bar Reef, Piankatank River, Virginia from May to October. Data from these two sites were averaged since there was no significant difference in temperature or salinity between sites (ANOVA, P < 0.05). Reference mean values for temperature and salinity data from 1993–95 are plotted with a solid line (\pm SE). Data from 1996 and 1997 are indicated by lines with symbols (\pm SE).

at the surface and just above the bottom with a Niskin bottle. Temperature was measured immediately with a thermometer and salinity was measured with a refractometer.

Finfish Sampling Periodicity and Methods.—Upper level trophic monitoring efforts focused on the pelagic finfish communities at each site. Fishes were classified based on mobility into one of three groups: resident, semi-resident, and transient. Resident fish species (all life history stages may be found in the habitat; Burchmore et al., 1985) are prey items for larger temporary or semi-resident species (fishes that use the habitat for a portion of their life cycle or on a seasonal basis; Burchmore et al., 1985; Chandler et al. 1985) as well as larger transient species (fishes that are found over a wide range of habitats; Burchmore et al., 1985).

Sampling efforts spanned seasonal, diurnal, and tidal scales. During 1996, regular sampling (gill netting, trawling, crab pot deployment and retrieval, nest substrate deployment and retrieval) was conducted bi-monthly on Palace Bar Reef from May through October during daylight hours and included at least two parts of the tidal cycle. Thirty-six hour sampling stations were conducted on Palace Bar Reef during the full moon in June and August 1996 incorporating tidal, diurnal, and seasonal variation. A total of 42 gill net sets,

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Table 1. Family, common, and scientific names with habitat use, life history stages, and species status for all fish species observed in the Piankatank River, Virginia during 1996-97. R denotes resident species, SR for semi-resident, and T for transient per Burchmore et al. (1985). Some semi-residential species may also be considered transient. Life history stages are classified as either adult (A) or juvenile (J) based on length (mm) and coloration. Gear type used to collect fishes include gill nets (G), trawls (T), crab pots (C), and nest substrates (N). Sampling methods and periodicity are detailed in the text

Family	Common name	Scientific name	Habitat use	Gear	Life	Recreationally or
					HistoryStage	Commercially valuable
						species
Anguillidae	American eel	Anguilla rostrata	ĸ	C, N, T	А	X
Batrachoididae	Oyster toadfish	Opsanus tau	R	C, N, T	A, J	
Blenniidae	Striped blenny	Chasmodes bosquianus	R	Z, T	A, J	
Bothidae	Summer flounder	Paralichthys dentatus	SR	G, T	A, J	X
Clupeidae	Blueback herring	Alosa aestivalis	Τ	ŋ	A	X
	Atlantic menhaden	Brevoortia tyrannus	Τ	Ŋ	A, J	X
Cyprinidae	Carp	Cyprinus carpio	Τ	Ð	Α	
Engraulidae	Bay anchovy	Anchoa mitchilli	Τ	Τ	Α	
Ephippidae	Spadefish	Chaetodipterus faber	SR	Τ	J	
Gadidae	Spotted hake	Urophycis regia	Τ	Τ	J	
Gobiesocidae	Skilletfish	Gobiesox strumosus	R	T, N	A, J	
Gobiidae	Naked goby	Gobiosoma bosc	ĸ	T, N	A, J	
Haemulidae	Pigfish	Orthopristis chrysoptera	Τ	C, T	А	
Labridae	Tautog	Tautoga onitis	SR	Z	J	X
Myliobatidae	Cownose ray	Rhinoptera bonasus	Τ	Ð	Α	
Percichthyidae	Striped bass	Morone saxatilis	Τ	G, T	A, J	X
Pomatomidae	Bluefish	Pomatomus saltatrix	Τ	C, G	А	×
Rachycentridae	Cobia	Rachycentron canadium	Τ	Ð	J	X
Sciaenidae	Atlantic croaker	Micropogonias undulatus	Τ	C, G, T	A, J	X
	Silver perch	Bairdiella chrysoura	Τ	G, T	A, J	
	Speckled trout	Cynoscion nebulosus	Τ	G, T	A, J	×
	Spot	Leiostomus xanthurus	Τ	G, T	A, J	
	Weakfish	Cynoscion regalis	T	G, T	A, J	X

Table 1. Continued.

Family	Common name	Scientific name	Habitat use	Gear	Life	Recreationally or
					HistoryStage	Commercially valuable
						species
Scombridae	Spanish mackerel	Scomberomorus maculatus	Τ	Ð	A	X
Serranidae	Black sea bass	Centropristis striata	SR	C	A	×
Soleidae	Hog choker	Trinectes maculatus	SR	G, T	A	
Sparidae	Pinfish	Lagodon rhomboides	SR	C, T	A	
Stromateidae	Butterfish	Peprilus triacanthus	Τ	G, T	A	
	Harvest fish	Peprilus alepidotus	Τ	Τ	J	
Syngnathidae	Lined seahorse	Hippocampus erectus	SR	Τ	A	
Tetraodontidae	Northern puffer	Sphoeroides maculatus	Τ	Τ	J	
Triglidae	Northern searobin	Priontus carolinus	R	Т	А	

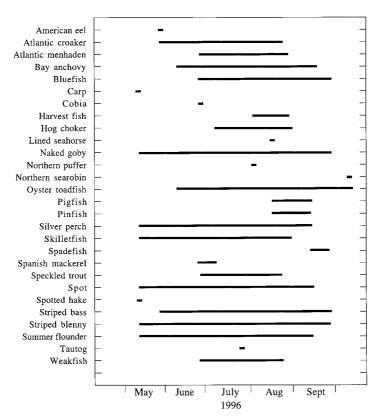


Figure 3. Fish species caught during 1996 with gill nets, otter trawls, crab pots, and nest substrates on Palace Bar Reef, Piankatank River, Virginia. Solid lines span the time interval during which a species was observed.

132 trawls, 120 crab pot sets, and 15 nest substrate deployments were completed during 1996. Sampling during 1997 relied exclusively on nine 36 h stations from May through September on both the new and full moon. On-reef and non-reef sites were sampled at 3 h intervals corresponding to changes in tidal stage. A total of 270 gill net sets, 172 crab pot sets, and 40 nest substrate deployments were completed during 1997.

Adults and juveniles of larger pelagic species were sampled using multi-panel experimental gill nets (one 30.5×1.8 m (1996 and 1997) and two 30.5×3.0 m (1997 only) nets all with one 7.6 m panel each of stretch square mesh monofilament of 57.2, 63.5, 73.0 and 76.2 mm). Gill nets were deployed such that the entire water column was sampled (e.g., the smallest net at the shallowest site) and were retrieved at 3 h intervals corresponding with changes in tidal stage. A 4.8 m otter trawl (25 mm stretch-mesh body, 7 mm mesh liner) was used in 1996 to sample benthic and smaller pelagic fishes on and immediately above the reef; all trawls were 5 min in duration at similar towing speeds (approximately 1 to 1.5 m s⁻¹) and were made above the reef and around the reef perimeter. Fishes were also collected in crab pots and benthic nest substrates. Nest substrates, benthic trays with potential nest sites (e.g., PVC tubes, paired valves of dead oysters, oyster shell), were deployed and checked bi-monthly throughout both 1996 and 1997 for related studies targeting naked gobies (*Gobiosoma bosc*) and striped blennies (*Chasmodes bosquianus*) (J. Harding, unpubl. data). Fishes were classified as "adult" or "juvenile" at the time of

Table 2. Fishes collected during 1997 with gill nets (G), crab pots (C), and nest substrates (N) in the Piankatank River, Virginia.

Fish species	Gear	Ginney	Roane	Palace Bar
•		Point	Point	Reef
American eel	N			X
Atlantic croaker	C, G	X	X	X
Atlantic menhaden	G	X	X	X
Black Sea bass	C	X		
Blueback herring	G	X	X	X
Bluefish	C, G	X	X	X
Butterfish	G			X
Carp	G		X	
Cownose ray	G			X
Hogchoker	G	X		X
Naked goby	N			X
Oyster toadfish	N			X
Silver perch	G	X	X	X
Skilletfish	N			X
Speckled trout	G	X	X	X
Spot	G	X	X	X
Striped bass	G	X	X	X
Striped blenny	N			X
Summer flounder	G	X		
Weakfish	G	X	X	X

collection using fork length (mm) (or total length in the absence of fork length) and body coloration.

Sampling efforts during 1996 focused on ichthyofauna observed on and immediately adjacent to Palace Bar Reef. Twenty-eight fish species from 24 families were caught on or immediately adjacent to Palace Bar Reef during 1996 (Table 1, Fig. 3). A total of 20 fish species representing 15 families were collected during 1997 across all three sites (Tables 1,2; Fig. 4). Resident benthic fishes including naked gobies, striped blennies, skilletfish, and oyster toadfish were regularly collected from Palace Bar Reef during both years. Transient fishes including bluefish, striped bass, Atlantic croaker, Atlantic menhaden, and spot were consistently caught on the reef as well as at non-reef locations during 1997 (Table 2; Fig. 3,4).

Data Analyses.—Significance levels for all statistical tests were established at P = 0.05 a priori. Bartlett's test for homogeneity of variance and the Ryan-Joiner test for normality were used prior to analyses (Minitab ver. 10.5). The reciprocal transformation was used for all count/abundance data per Zar (1996) prior to analyses to satisfy assumptions of homogeneity of variance and normality.

Abundance of spat, small oysters, and market oysters on Palace Bar and Palace Bar Reef from 1996 and 1997 was compared using a two-factor ANOVA (year \times site); data were transformed with the reciprocal transformation (Zar, 1996). Fisher's test was used for post-hoc multiple comparisons. Spat abundance was similar between sites in both years (Table 3; ANOVA, P > 0.05). Market oyster abundance was higher on Palace Bar in 1996 (ANOVA, P < 0.05) but similar between sites in 1997 (Table 3; ANOVA, P > 0.05). Small oyster abundance was greater on Palace Bar during 1996 and greater on Palace

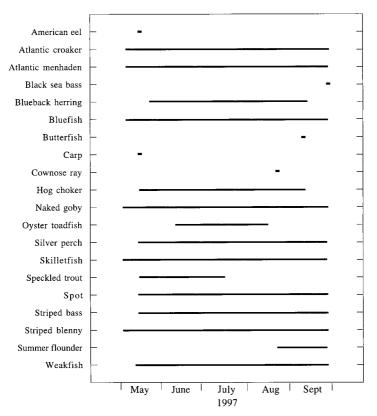


Figure 4. Fish species caught during 1997 with gill nets, crab pots, and nest substrates from Ginney Point, Roane Point, and Palace Bar Reef, Piankatank River, Virginia. Solid lines span the time interval during which a species was observed.

Bar Reef in 1997 (Table 3).

Water temperature and salinity data were transformed (natural logarithm) prior to analyses and satisfied assumptions of both homogeneity of variance and normality. Neither water temperatures nor salinity values were significantly different between depths (surface or near-bottom) or sampling sites (ANOVA, P < 0.05). Salinities, but not water temperatures, were significantly different between years: 1996 and 1997 (ANOVA, P < 0.05). Recorded weekly water temperatures during 1996 and 1997 were similar to those observed weekly at the same sites during 1993–95 (Fig. 2, R. Mann, unpubl. data). Salinities observed during 1996 were the lowest observed at these sites from 1993–97.

Total number of fish species collected with gill nets at Palace Bar Reef were compared between 1996 and 1997 using an ANOVA to assess potential interannual variation at the reef site. Data were transformed (reciprocal transformation, per Zar, 1996) prior to analyses and satisfied assumptions of both homogeneity of variance and normality. There was no significant difference between numbers of species caught in gill nets at Palace Bar Reef between years (ANOVA, P > 0.05).

Numbers of fish species caught with gill nets during the 1997 36-h sampling stations were compared between sites and across seasons using an ANOVA with site and day of the year as factors. The reciprocal transformation (Zar, 1996) was used and data satisfied assumptions of both homogeneity of variance and normality. Neither site nor season (as

Table 3. Comparision of oyster abundance on Palace Bar and Palace Bar Reef, Piankatank River, Viginia during 1996 and 1997. Average counts of spat, small, and market oysters m⁻² are presented with standard error (SE); n refers to the number of samples collected. "Spat" refers to young-of-the-year oysters; "small" oysters are less than 7.6 cm (maximum dimension) while "market" oysters are greater than 7.6 cm (maximum dimension). Sampling methods are detailed in the text. Vertical lines are to the right of samples that are statistically similar at the P = 0.05 level (ANOVA).

Year	Site	n	Average No. market oysters m ⁻² (± SE)	Average No. small oysters m ⁻² (± SE)	Average No. spat m ⁻² (± SE)
1996	Palace Bar	4	2.0 (1.4)	165.0 (30.4)	22.5 (2.6)
	Palace Bar Reef	30	0 (0)	55.7 (4.2)	24.0 (2.9)
1997	Palace Bar	4	3.5 (2.1)	19.0 (6.2)	77.5 (23.0)
	Palace Bar Reef	12	1.8 (1.0)	43.6 (5.0)	4.9 (1.5)

indicated by day of the year) significantly affected the numbers of fish species caught with gill nets (ANOVA, P > 0.05).

DISCUSSION

Palace Bar Reef, Piankatank River, Virginia supports a diverse fish fauna; 32 species of finfishes representing 26 families were observed on or in proximity to Palace Bar Reef during 1996–97. Differences in sampling gear, gear efficiency, and sampling effort between years may account for differences in fish species richness (20 species vs 28) between years. Many of the fish species observed in relation to Palace Bar Reef during 1996 but not 1997 were captured with trawls, e.g., bay anchovies, northern searobins, spadefish, spotted hake. Presence of benthic and midwater species smaller than gill net meshes at Roane Point and Ginney Point are unknown since these sites were sampled with gill nets only. Gill netting data showed no differences between these three sites, but it is possible that there would be differences between sites in benthic and midwater fish species because of the absence of relief (both Ginney Point and Roane Point) and substrate heterogeneity due to shells (Roane Point). Further study is needed to clarify habitat use patterns by these fishes.

Given that the Piankatank River is a small Chesapeake Bay subestuary, high levels of species richness might be surprising unless placed in context or associated with living oyster reefs. Other studies on natural or restored oyster reefs within Chesapeake Bay have observed similarly high numbers of finfish species. Breitburg (1998) describes 17 fish species representing 14 families at Flag Pond oyster reef, near the Patuxent River, Maryland (salinity range 10 to 17‰, water temperature range 19 to 29°C). Nestlerode et al. (in prep.) reports 34 fish species from 24 families on restored oyster reefs near Fisherman's Island, Virginia (salinity range 33–35 ‰, water temperature range 15–28°C). Natural and restored oyster reefs within Chesapeake Bay support diverse fish assemblages along a gradient of physical conditions. By comparison, Feigenbaum et al. (1989) report nine fish species in relation to an artificial "fishing" reef near Gwynn's Island (approximately 12 km east of Palace Bar Reef in the mainstem of Chesapeake Bay).

Differences in sampling regimes among sites and studies make quantitative comparisons between species assemblages difficult; standardization of sampling methods and effort would reduce both seasonal and gear biases. However, on the basis of available presence/ absence data, natural oyster reefs within Chesapeake Bay support high numbers of fish species. Species richness in and of itself may be an indicator of relative community rehabilitation or trophic health. Transient species are unlikely to be consistently present within a habitat, especially a small subestuary, unless they are deriving some benefit. In this case, many of the transient species observed on and around Palace Bar Reef were foraging on the reef (Mann and Harding, 1997; Harding and Mann, in review; J. Harding, unpubl. data). Interannual species-specific diet and abundance patterns by site, season, and tidal stage for transient fishes caught in the Piankatank River by site are currently under investigation (J. Harding and R. Mann, unpubl. data).

Increased levels of fish species richness associated with oyster reefs is not surprising in light of the trophic interconnections between the oysters themselves and resident, semi-resident, and transient fishes. Success of lower trophic levels enhances production at the upper levels; living reefs host a strong trophic network and have high potential for long term productivity because the resources necessary for trophic enhancement are inherent in the habitat. As reef restoration efforts continue in estuarine habitats, the trophic impacts of reef restoration have ramifications for recreationally and commercially valuable finfish communities.

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