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# 2010 Annual Report Estimating Relative Juvenile Abundance of Ecologically Important Finfish in the Virginia Portion of Chesapeake Bay Project Number: F-104-R-14 

Troy D. Tuckey<br>Virginia Institute of Marine Science<br>Mary C. Fabrizio<br>Virginia Institute of Marine Science

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# Estimating Relative Juvenile Abundance of Ecologically Important Finfish in the Virginia Portion of Chesapeake Bay 

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Cover: Juvenile spot captured by trawl.

DISCLAIMER
Some of the results contained in this report have recently been completed and may contain errors and/or need further refinement.

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## EXECUTIVE SUMMARY

June 2009 - May 2010
The juvenile fisheries trawl survey conducted by the Virginia Institute of Marine Science (VIMS) is the oldest continuing monitoring program (55 years) for marine and estuarine fishes in the United States. This survey provides a monthly assessment of abundance of juvenile marine and estuarine fishes and crustaceans in the tidal rivers and main stem of Chesapeake Bay. The survey provides crucial data to state, regional, and national fisheries management agencies, including the Virginia Marine Resources Commission (VMRC), the Atlantic States Marine Fisheries Commission (ASMFC), the Mid-Atlantic Fisheries Management Council (MAFMC), and the National Marine Fisheries Service (NMFS). The MAFMC recognizes the VIMS Juvenile Trawl Survey as one of the key predictors of summer flounder recruitment, and the American eel index was a vital component of the assessment that led to the 2006 ASMFC American Eel Management Plan.

Several annual indices of juvenile abundance have been generated from Trawl Survey data for species of key recreational, ecological, and commercial importance in the Virginia portion of Chesapeake Bay. These include spot, Atlantic croaker, weakfish, summer flounder, black sea bass, scup, striped bass, white perch, white catfish, channel catfish, blue catfish, silver perch, American eel, and bay anchovy. Historically, four different estimates of relative abundance were developed and reported for juvenile finfish in the survey. However, only the unconverted indices (Random Stratified Index - RSI, 1988 to present) for the target species are the focus of this report. We chose to use this index because it is based on data collected from a random stratified survey design with consistent spatial and temporal domains. Furthermore, gear changes since 1988 were minor and inconsequential (replacement of standard trawl doors with China-V doors in 1991), thus, the index can be calculated without the use of gear conversion factors.

In 2009, abundance indices for American eel, black sea bass, channel and white catfish (YOY and age $1+$ ), spot, striped bass, summer flounder, and white perch age $1+$ were below the time series average (1988-2009, based on the RSI). Abundance indices for YOY Atlantic croaker, bay anchovy, blue catfish age $1+$, scup, and silver perch indicated above average recruitment, while YOY blue catfish, weakfish, and white perch experienced average recruitment.

In May 2010 a scale model of the trawl survey gear was tested in a flume tank at Memorial University, Centre for Sustained Aquatic Resources, Newfoundland, Canada. Results from the gear testing project will allow us to optimize the attachment of gear mensuration sensors to the trawl net to measure the wing spread without affecting how the trawl net currently fishes. With the measured wing spread and GPS starting and ending positions, relative abundance data can be calculated on an area swept basis instead of assuming a constant area is sampled by each tow as is the current practice.

Routine sampling in Mobjack Bay, located on the western side of Chesapeake Bay and north of the York River entrance, was initiated in March 2010. Mobjack Bay sampling stations consist of a mix of seven fixed and ten random stratified stations each month that encompass the main portion of the bay and the four river systems that empty into it. Extensive seagrass beds occur in Mobjack bay and likely serve as important nursery habitat for juvenile fishes. Results from Mobjack Bay sampling were analyzed separately to maintain consistency with spatial coverage discussed in previous reports. In three months of sampling ( 51 stations), 26 species were captured totaling 34,593 individuals. Eleven species were captured in greater relative abundance in Mobjack Bay compared with a subset of data for the same time period (March - May) from the full trawl survey.

## INTRODUCTION

Relative abundance estimates of early juvenile (age 0 ) fishes and invertebrates generated from fishery-independent survey programs provide a reliable and early indicator of year-class strength (Goodyear, 1985), and may be used to evaluate the efficacy of management actions. The Chesapeake Bay Stock Assessment Committee (CBSAC) reviewed available indices of juvenile abundance for important fishery resources in Chesapeake Bay (hereafter referred to as "Bay") and recommended that "a unified, consistent trawl program should be one of the primary monitoring tools for finfish and crab stock assessment" (Chesapeake Bay Program Stock Assessment Plan, Chesapeake Executive Council, 1988). Subsequently, pilot studies directed at developing a comprehensive trawl survey for Chesapeake Bay began at VIMS with monthly trawl sampling in the main stem of the lower Bay. This effort complemented and expanded the monthly trawl sampling conducted in major Virginia tributaries (James, York, and Rappahannock rivers) by the Virginia Institute of Marine Science (VIMS).

The present sampling program, which includes the Bay and its tributaries, ensures that data are of sufficient geographic coverage to generate relative abundance indices for recreationally, commercially, and ecologically important finfishes and invertebrates. The National Marine Fisheries Service Marine Recreational Fisheries Statistics Survey shows that recreational catches in Virginia are dominated by Atlantic croaker (Micropogonias undulatus), summer flounder (Paralichthys dentatus), spot (Leiostomus xanthurus), striped bass (Morone saxatilis), black sea bass (Centropristis striata), bluefish (Pomatomus saltatrix), pigfish (Orthopristis chrysoptera), weakfish (Cynoscion regalis), and kingfishes (Menticirrhus spp.). These species depend on the lower Chesapeake Bay and its tributaries as nursery areas and, with the exception of bluefish, are highly vulnerable to bottom trawls. Additional species of recreational interest, such as scup (Stenotomus chrysops), white perch
(Morone americana), silver perch (Bairdiella chrysoura), white catfish (Ameiurus catus), channel catfish (Ictalurus punctatus) and blue catfish (I. furcatus), are also taken with sufficient regularity during trawling operations to provide information suitable for the generation of juvenile abundance indices. Although annual juvenile indices are the primary focus of this project, survey results can be used to address other aspects of finfish population biology, such as habitat utilization, early growth and survival, environmental effects on recruitment, or disease prevalence. For example, episodic climatic events, such as hurricanes, affect recruitment of shelf spawning species such as Atlantic croaker (Montane and Austin, 2005).

The development of juvenile indices requires a continuous time series of data to determine the most appropriate area-time sequences for index calculations. Provisional annual juvenile abundance indices were developed for spot, weakfish, Atlantic croaker, summer flounder, and black sea bass (Colvocoresses and Geer, 1991), followed by scup (Colvocoresses et al., 1992), white perch and striped bass (Geer et al., 1994), and white catfish, channel catfish, and silver perch (Geer and Austin, 1994). More recently, blue catfish, American eel (Anguilla rostrata), and bay anchovy (Anchoa mitchilli) indices were developed. Through the use of gear conversion factors and post stratification, a time series of index values can be produced back to 1955 for most species (Geer and Austin, 1997).

Many species of interest are captured in significant numbers across several year classes. As a result, both juvenile and age $1+$ (i.e., all fish older than age 0 ) indices are reported for white perch, white catfish, channel catfish, and blue catfish.

This report summarizes the activity of the VIMS Juvenile Finfish Trawl Survey from June 2009 through May 2010. Abundance indices are provided from 1988 to the present, along with the mean value estimated across the time series; indices for years prior to 1988 are available in previous reports.

## METHODS

Field Sampling

The field sampling protocol is described in detail in Lowery and Geer (2000). In brief, a 30' $(9.14 \mathrm{~m})$ semi-balloon otter trawl, with $1.5^{\prime \prime}(38.1 \mathrm{~mm})$ stretched mesh and $0.25^{\prime \prime}(6.35 \mathrm{~mm})$ cod-end liner, is towed along the bottom for 5 minutes during daylight hours. Sampling in the Bay occurs monthly except during January and March, when few target species are available. Sampling in the tributaries also occurs monthly, at both the random stratified and historical fixed (mid-channel) stations. The stratification system is based on depth and latitudinal regions in the Bay, or depth and longitudinal regions in the rivers. Each Bay region spans 15 latitudinal minutes and consists of six strata: western and eastern shore shallow ( $4-12 \mathrm{ft}$ ), western and eastern shoal ( $12-30 \mathrm{ft}$ ), central plain (30-42 ft), and deep channel ( $\geq 42 \mathrm{ft}$ ). Each tributary is partitioned into four regions of approximately ten longitudinal minutes, with four depth strata in each $(4-12 \mathrm{ft}, 12-30 \mathrm{ft}, 30-42 \mathrm{ft}$, and $\geq 42 \mathrm{ft}$; Figure 1). Strata are collapsed in areas where certain depths are limited. Fixed stations were assigned to a stratum according to their location and depth.

With the exception of the fixed river stations, trawling sites within strata are selected randomly from the National Ocean Service's Chesapeake Bay bathymetric grid, a database of depth records measured or calculated at 15-cartographic-second intervals. Between two and four trawling sites are randomly selected for each Bay stratum each month, and the number varies seasonally. Exceptions include the shallow water strata where only a single station is sampled each month. For most river strata, one to two random stations are selected per month. Sampling in the York River has been altered slightly as of 1991 to make depth strata similar to those in the James and Rappahannock rivers and main stem Bay. The stratification scheme for the tributaries was modified in January 1996 to create separate depth strata of $30-42 \mathrm{ft}$ and $\geq 42 \mathrm{ft}$ (Geer and Austin, 1996). Because tributary
sampling had occurred at these depths prior to 1996, samples collected previously were reassigned to the new strata established in 1996.

Fixed stations were sampled monthly (nearly continuously) since 1980 with sites in each tributary spaced at approximately 5-mile intervals from the river mouth up to the freshwater interface. From the mid-1950's (York River) and early-1960's (James and Rappahannock rivers) to 1972, fixed stations were sampled monthly using an unlined $30^{\prime}$ trawl (gear code 010). During 1973-79, semiannual random stratified sampling was performed by the VIMS Ichthyology Department, while the VIMS Crustaceology Department continued monitoring the fixed tributary stations on a limited monthly basis (May - November). Area-based weightings for the tributaries were previously assigned by dividing each river into two approximately equal length 'strata' by assuming that the stations in each stratum were representative of the channel areas in those reaches (see Lowery and Geer, 2000). As of 1996, all three tributaries were sampled with a random stratified design; the fixed stations were assigned to a stratum based on location and depth. The current design (combined fixed and random stations) provides greater spatial coverage and a long-term historical reference.

At the completion of each tow, all fishes are identified to species, counted, and measured to the nearest millimeter fork length (FL), total length (TL), or total length centerline (TLC, black sea bass only). Species that have varying size ranges are measured and counted by size class and large catches of a particular species are randomly subsampled, measured, and the remaining unmeasured catch is counted. In instances of extremely large catches (e.g., bay anchovy), subsampling is performed volumetrically.

Since May 1998, habitat or substrate type sampled by the trawl has been recorded (Table 1). Fish distribution and abundance may be influenced by various substrates such as shell, sponge, hydroids, and sea squirts that may be used as shelter, spawning habitat or for feeding. Substrates are
measured at each trawling site based on the quantity (volume in a standard container) observed in the net. Volumetric measurements of gelatinous zooplankton are also recorded for each trawl site because large catches of jellyfish and ctenophores may affect the catch (e.g., changes in gear saturation or efficiency).

## Juvenile Index Computations

Many of the target species of this study are migratory and abundance estimation presents special difficulties, particularly if the timing and duration of migration varies annually. Juvenile fishes that use estuarine nursery areas are especially vulnerable to the vagaries of the environment, as many rely on wind-driven and tidal circulation patterns for transport into the estuaries as larvae and early juveniles (Norcross, 1983; Bodolus, 1994; Wood, 2000). The outward migration of some species from the nursery area may follow annually variable environmental cues (e.g., temperature changes). Ideally, juvenile abundance should be measured when young fish are fully recruited to the nursery area under study. In practice, however, this can only be accomplished if the time of maximal abundance and size at recruitment to the gear can be predicted (and surveys timed accordingly), or if surveys can be conducted with high frequency over the season of potential maximal abundance. Neither of these two approaches is practical. The period of maximal abundance and the scope of the area occupied by juvenile fish have proven to be variable among years and among species. This observation, coupled with multi-species monitoring objectives, precludes temporally intense surveys. Consequently, the survey is operated with a regular periodicity (monthly) and sample-site selection is performed using a standard sampling design for multispecies resource surveys.

A monthly size threshold value is applied to the length frequency information collected for each target species to partition the catch data into young-of-year and older components for juvenile index calculation (Table 2). Threshold values vary among months for each species and are based on
modal analyses of historical, composite length-frequency data and on reviews of ageing studies (Colvocoresses and Geer, 1991). For earlier months of the biological year, threshold values are usually arbitrary and fall between completely discrete modal size ranges. In the later part of the biological year, when the size of early spawned, rapidly growing individuals of the most recent year class may approach that of later spawned, slower growing individuals of the previous year class, threshold values were selected to preserve the numeric proportionality between year classes despite the potential misclassification of some individuals (Table 2). The extent of overlapping lengths and the proportion within that range attributable to each year class was estimated based on the shapes of the modal curve during the months prior to the occurrence of overlap. A length value was then selected, which preserves the proportional separation of year classes. Although this process involved considerable subjectivity and ignored possible interannual variability in average growth rates, the likelihood of significant error is small, because only a small fraction of the total number of young-ofyear individuals falls within the zone of overlap, and furthermore, most of the data used to construct juvenile indices were drawn from months when no overlap was present.

After removing non young-of-year individuals, monthly stratum-specific abundances and occurrence rates are calculated for each target species. Numbers of individuals captured are log transformed $(\ln (\mathrm{n}+1))$ prior to abundance calculations following Chittenden (1991). Average catch rates (and the approximate $95 \%$ confidence intervals as estimated by $\pm 2$ standard errors of the mean) are then back-transformed to geometric means. The stratum-specific coefficient of variation is expressed as the standard deviation divided by the log-transformed mean catch: STD/ $\mathrm{EY}_{\text {st }}$ (Cochran, 1977). The catch data were examined for area-time combinations that provided the best basis for juvenile index calculations. Criteria applied during the selection process included identification of maximal abundance levels, uniformity of distribution, minimization of overall variance, and
avoidance of periods in which distribution patterns indicated migratory behavior. Although identification of areas most suitable for index calculations (primary nursery zones) was generally clear, selection of appropriate time windows was more complex. Surveys are timed on regular monthly intervals that may or may not coincide with periods of maximal recruitment to the nursery areas. The use of a single (maximal) month's survey result is therefore inappropriate and would decrease sample size, increase confidence intervals, and increase the risk of sampling artifacts. Conversely, the temporal series of data incorporated into index calculations should not be longer than necessary to capture the period of maximal juvenile use of the nursery area. With this approach, three- or four-month periods (6 months for bay anchovies) that provided reasonable abundance data for each species were identified (Table 2).

Using these catch data from area-time combinations, an annual juvenile index is currently calculated as the weighted geometric mean catch per tow. This is accomplished by calculating stratum-specific means and variances and combining the stratum-specific estimates using weights based on stratum area (Cochran, 1977). Because stratum areas are not uniform, a weighted mean provides an index that more closely approximates actual population abundance.

The following three indices are produced for each species for 1988 to the present: an index based on the current Bay strata plus the fixed mid-channel tributary stations (Bay \& River Index BRI), an index based on the fixed mid-channel tributary stations only (River Only - RO), and a random stratified index using all spatially appropriate data (Random Stratified Index - RSI; in previous reports, this index was called the Random Stratified Converted Index, RSCI). Data collected prior to 1988 are excluded from this report because results from the longer time series are considered provisional (i.e., indices prior to 1988 require both gear and vessel conversion factors, and
concerns about conversion factors for this period are being addressed). Multiple indices are presented in this report for completeness, but usually only the RSI will be described in detail.

## Mobjack Bay

Routine sampling in Mobjack Bay, located on the western side of Chesapeake Bay and north of the York River entrance, was initiated in March 2010 (Figure 1). Extensive seagrass beds occur in Mobjack Bay and likely serve as important nursery habitat for juvenile fishes. Seventeen stations were sampled each month (seven fixed stations and ten random stations) following the depth stratification scheme decribed previously for the trawl survey. Three fixed stations occur in the main portion of Mobjack Bay, one fixed station occurs in each river (Ware, East, Severn, and North rivers), and the ten depth-stratified random stations are selected from sites located throughout the bay and river system.

## RESULTS

A summary of tows completed from 1988 through May 2010 (Table 3) provides a comprehensive synopsis of the sampling completed for this report. For the 2009-2010 project year (June through May), 1224 sites were sampled resulting in 562,635 fishes identified and enumerated from 105 different species (Table 4). Bay anchovy, hogchoker, and Atlantic croaker accounted for $84.4 \%$ of the catch by numbers. Ignoring bay anchovy and hogchoker, five species - Atlantic croaker, spot, weakfish, white perch, and blue catfish - represented $77.8 \%$ of the catch numerically (Table 4).

Indices were calculated and described for the following species: American eel, Atlantic croaker, bay anchovy, black sea bass, blue catfish, channel catfish, scup, silver perch, striped bass, spot, summer flounder, weakfish, white catfish, and white perch. Length frequency distributions for
each species are shown in Appendix Figure 1 with index-sized fish indicated in red. Actual relative abundance indices are calculated on a subset of the data as described below.

American eel (Anguilla rostrata) - American eel are present along the Atlantic and Gulf coasts of North America and inland in the St. Lawrence Seaway and Great Lakes (Murdy et al., 1997). This catadromous species is panmictic and supported throughout its range by a single spawning population (Haro et al., 2000). Spawning takes place during winter to early spring in the Sargasso Sea. The eggs hatch into leaf-shaped, ribbon-like larvae called leptocephali, which are transported by ocean currents (over 9-12 months) in a generally northwesterly direction. Within a year, metamorphosis into the next life stage (glass eel) occurs in the Western Atlantic near the east coast of North America. Coastal currents and active migration transport the glass eels into rivers and estuaries from February to June in Virginia and Maryland. As growth continues, eels become pigmented (elver stage) and within $12-14$ months, eels acquire a dark color with underlying yellow (yellow eel stage). Many eels migrate upriver into freshwater rivers, streams, lakes, and ponds, while others remain in estuaries. Most of the eel's life is spent in these habitats as a yellow eel. Age at maturity varies greatly with location and latitude, and in Chesapeake Bay may range from 8 to 24 years, with most eels in the Bay area less than 7 years old (Owens and Geer, 2003). Eels from Chesapeake Bay mature and migrate at an earlier age than eels from northern areas (Hedgepeth, 1983). Metamorphosis into the silver eel stage occurs during the seaward migration that occurs from late summer through autumn, as mature eels migrate back to the Sargasso Sea to spawn and die (Haro et al., 2000).

The current American eel index includes all eels ( $>152 \mathrm{~mm} \mathrm{TL}$ ) collected in the upper half of each of the major tributaries (Figure 1, JA 3 and 4, YK 3 and 4, and RA 3 and 4) during April, May, and June. American eel indices exhibited above average recruitment (mean RSI $=2.13$ ) in the late

1980's and early 1990's and below average recruitment thereafter including 2009 (Table 5; Figure 2, top). American eel are more abundant in the Rappahannock River compared with the James and York rivers (Figure 2, bottom).

Atlantic croaker (Micropogonias undulatus) - Atlantic croaker are typically captured in high abundance and are widely distributed throughout the survey area (Figure 3, bottom). Spawning takes place over a protracted period, such that small juveniles ( $<30 \mathrm{~mm} \mathrm{TL}$ ) can be present in catches yearround (Norcross, 1983; Colvocoresses and Geer, 1991; Colvocoresses et al., 1992; Geer et al., 1994). For some year classes, peak abundance occurs in the fall at lengths less than 100 mm TL , but for other year classes, the peak occurs the following spring. Previously, we provided two estimates of the index: a juvenile fall index (October - December) based on catches in the tributaries, and a spring recruit index (May - August) based on catches in the Bay and tributaries combined. Because the fall index does not reflect over-winter mortality, only the spring index is presented. The Atlantic croaker spring RSI has remained stable throughout the time series with recruitment in 2009 above the historic average (Table 6 and Figure 3, top; mean RSI $=1.26$ ).

Bay anchovy (Anchoa mitchilli) - Bay anchovy are the most abundant finfish throughout Chesapeake Bay and its tributaries, and are found in salinities ranging from 1-33 \% (Murdy et al., 1997). Bay anchovy feed mostly on zooplankton and are an important prey of other Bay fishes (Murdy et al., 1997). In years of average freshwater inflow (e.g.,1997-2000), Atlantic menhaden, bay anchovy, and Atlantic croaker often dominate fish biomass in Chesapeake Bay (Jung, 2002). Bay anchovy abundance has increased in recent years from a period of low recruitment observed during 2001-2002 (Table 7; Figure 4, top). The 2009 index was above the time-series average (mean RSI $=$ 24.32). As expected, bay anchovy are ubiquitous in trawl survey catches (Figure 4, bottom).

Black sea bass (Centropristis striata) - Black sea bass are seldom taken in large numbers but regularly occur in survey catches. Young-of-year black sea bass occur throughout the Bay and appear occasionally in the lower portions of the tributaries (Figure 5, bottom). Index calculations are based on all Bay strata and the lower James stratum from May through July. Although some early juveniles appear in the Bay during their first summer and fall, more young-of-year enter the estuary during the following spring. Black sea bass spawn in the summer in the Mid-Atlantic Bight (Musick and Mercer, 1977). Thus, the index is calculated for the year class spawned the previous calendar year (i.e., the index for the 2008 year class is based on catches from May to July 2009). The black sea bass RSI was generally above average (mean $\mathrm{RSI}=0.72$ ) prior to 1995 , but fell below average the following years with the exception of 2001 and 2007 (Table 8; Figure 5, top).

Blue catfish (Ictalurus furcatus) - The blue catfish, one of Virginia's largest freshwater fishes (Jenkins and Burkhead, 1993), was introduced to the Chesapeake Bay area as a sportfish in the James, Rappahannock, and Mattaponi rivers between 1974 and 1989 (Virginia Department of Game and Inland Fisheries, 1989 as reported by Connelly, 2001). The blue catfish is a carnivorous bottom feeder that inhabits the main channels and backwaters of rivers (Murdy et al., 1997). Blue catfish are collected from the mesohaline portions of the major tributaries upstream to the limits of the trawl survey (Figures 6 and 7, bottom). The juvenile blue catfish RSI exhibited an increasing trend over time with above average recruitment (mean $\operatorname{RSI}=0.74$ ) in 1996, 1997, 2003 - 2006, after which the index declined (Table 9; Figure 6, top). The abundance index for age $1+$ blue catfish RSI has been above average (mean RSI $=1.82$ ) since 2004 (Table 10; Figure 7, top).

Blue catfish indices have increased since 1988 and the ecosystem effects of this introduced species are unknown. Diets of small blue catfish are dominated by invertebrates (mostly amphipods, isopods and mud crabs), while larger blue catfish diets include invertebrates, menhaden (Brevoortia
tyrannus), and gizzard shad (Dorosoma cepedianum; Parthree et al., 2008). Other catfishes (white and channel) have similar diets and may be competing with the introduced blue catfish for the same prey resources.

Channel catfish (Ictalurus punctatus) and White catfish (Ameiurus catus) - Channel catfish and white catfish are usually found in the upper portions of the tributaries (Figures 8, 9, 16, and 17, bottom). Although each river system is unique, spawning typically occurs in late May through early July in Virginia (Menzel, 1945); consequently, June was selected as the start of the biological year. The survey typically catches both species up to 600 mm FL with juveniles ( $\leq 50 \mathrm{~mm}$ FL) first recruiting to the gear in June. In most years, juvenile recruitment occurs from January to April for both species in the upriver strata only.

The channel catfish was introduced to Virginia in the late 1800s (Jenkins and Burkhead, 1993). Juvenile channel catfish exhibited low recruitment in most years with a few notable peaks, and in 2009, one juvenile channel catfish was captured by the VIMS trawl survey. The age 1+ channel catfish RSI has exhibited below average (mean $R S I=0.84$ ) values since 1997 (Tables 11 and 12; Figures 8 and 9, top). Similarly, RSIs for juvenile white catfish indicate below average recruitment (mean $\mathrm{RSI}=0.38$ ) since 1994 and age $1+$ white catfish exhibited below average abundance (mean RSI = 1.25) since 1998 (Tables 19 and 20; Figures 16 and 17, top).

Scup (Stenotomus chrysops) - Scup are primarily a marine, summer spawning species that use the Bay in a manner similar to black sea bass. The estuary is rarely used as a nursery area by early juveniles, but older juveniles can be found in the Bay during their second summer. Early juvenile scup (25-40 mm FL) occasionally appear in survey catches in June. Older scup first appear in catches in May, and by June, they range from 50 to 215 mm FL. Using the original length threshold for scup that was based on ageing studies (Morse, 1978), trawl survey catches were found
to typically include three age groups (age 0 , age 1 , and age $2+$ ). Because catches of age 0 and age $2+$ scup are highly variable and low, index calculations using trawl survey catches are based on age 1 individuals only. Age 1 fish are present in the Bay and available to the gear for the entire summer and early fall.

Scup are predominantly collected in the lower Bay (Figure 10, bottom). Catch rates for scup usually peak in July, and the index is calculated from catches taken in June to September. Scup indices have increased in recent years with the 2006 index approaching the maximum observed since 1988 (mean RSI = 1.37; Table 13; Figure 10, top).

Silver perch (Bairdiella chrysoura) - Silver perch are found in all sampling strata (Figure 11, bottom). Spawning occurs in the deep waters of the Bay and offshore from May to July, and juveniles ( $\leq 100 \mathrm{~mm}$ TL) enter the Bay by July (Chao and Musick, 1977; Rhodes, 1971). Abundance indices for silver perch from the RSI are consistent and stable (Table 14; Figure 11, top). The time series average RSI index for silver perch is 0.61 .

Spot (Leiostomus xanthurus) - Spot indices are calculated using all strata from July to October. Spot is often one of the most abundant recreational species captured by the survey, however compared with catches in the late 1980s and early 1990s, numbers have declined (Table 15; Figure 12, top). The RSI index in 2009 was below average for the time series (mean RSI = 12.42). Spot are widely distributed throughout the Bay and tributaries (Figure 12, bottom).

Striped bass (Morone saxatilis) - Striped bass use the upper tributaries of the Bay as spawning and nursery grounds; spawning occurs from early to mid-April through the end of May, in tidal freshwater areas just above the salt wedge. Juvenile striped bass often appear in catches from May to July in size classes less than 50 to 100 mm FL (Table 16) during years of greater abundance, but then diminish in abundance until the following winter. A second, stronger and more consistent
period of juvenile abundance occurs in December and continues through February in the upper regions of the rivers. The trawl survey index for striped bass is based on this winter recruitment period. Collections of juvenile striped bass occurred in the major tributaries only (Figure 13, bottom).

Juvenile striped bass showed strong recruitment peaks for the 1993 and 2000 year classes but recruitment has declined in recent years (Table 16; Figure 13, top). Index values for the past 9 years have been below the time-series average (mean $\mathrm{RSI}=0.87$ ).

Summer flounder (Paralichthys dentatus) - Summer flounder spawn on the continental shelf from September through January with the peak occurring in October and November (Murdy et al., 1997). Flounder larvae enter the Bay and other Virginia estuaries from October through May with juveniles using shallow fine-substrate areas adjacent to seagrass beds (Murdy et al., 1997; Wyanski, 1990). Low water temperatures can have significant effects on growth and survival of individuals that enter the estuary in the winter (Able and Fahay, 1998). Juvenile summer flounder first appear in catches in late March, which is used as the beginning of the biological year. Juvenile summer flounder abundance continues to increase steadily throughout the summer and early fall to a late fall peak, and then trawl catches decline, presumably reflecting emigration of young fish during December. For our trawl survey, September, October, and November usually encompass the months of greatest abundance of juvenile summer flounder. Juvenile flounder are broadly distributed throughout the Bay and lower rivers. Consequently, index calculations are based on catches from the Bay and lower river strata during September, October, and November.

Juvenile summer flounder indices were greater during the early 1990s compared with recent years, but recruitment appears to be consistent since 1995 (Table 17; Figure 14, top). The 2009 index
was below average (mean $\mathrm{RSI}=1.03$ ). Juvenile summer flounder were captured throughout the Bay and lower portions of the rivers, particularly the York River (Figure 14, bottom).

Weakfish (Cynoscion regalis) - Weakfish are one of the dominant species in our catches, and juveniles are found throughout the Bay and tributaries (Figure 15, bottom). Juveniles have occurred in catches in late May and June, with June considered the beginning of the biological year. The weakfish index (RSI) is consistent and indicates steady recruitment since 1988 (mean RSI $=6.48$; Table 18; Figure 15, top).

White perch (Morone americana) - Spawning of white perch occurs in the upper tributaries from March to July with a peak occurring from late April to early May. Index months include December to February for juveniles and November to February for age 1+. Index stations are from the upper river strata.

Juvenile (age 0 ) white perch recruitment has been variable with a time series average RSI of 8.22, whereas abundance of age $1+$ white perch has been declining (Tables 21 and 22; Figures 18 and 19, top). Abundance of age $1+$ white perch has been below average (mean RSI $=14.88$ ) in 15 of the last 18 years. White perch are collected throughout the rivers, but are most abundant in the upper river stations (Figure 18 and 19, bottom).

## Mobjack Bay

Routine sampling in Mobjack Bay began in March 2010 and a total of 34,593 fishes were collected in 51 tows (Appendix Table 2). Twenty-six species were captured with an average of 678 individuals per tow. Compared with the full trawl survey from the same time period (March - May), Mobjack Bay produced 1.8 times as many fish per tow. Eleven species were captured in greater relative abundance compared with the full survey from March to May: bay anchovy, spot, weakfish,

Atlantic herring, northern pipefish, Atlantic cutlassfish, butterfish, naked goby, lined seahorse, feather blenny, and rough scad.

## DISCUSSION

Juvenile indices contribute to the assessment and management of important recreational and commercial species in Chesapeake Bay and the mid-Atlantic Bight. For example, the VIMS Trawl Survey was recognized by the Mid-Atlantic Fisheries Management Council (MAFMC) as an important source of the summer flounder recruitment index; the VIMS index was instrumental in shaping more protective harvest regulations in Virginia. Other indices utilized by management agencies include those for Atlantic croaker, spot, and weakfish. Though the trawl is not the preferred gear with which to sample American eel, eel indices from the trawl survey played an important role in the 2006 ASMFC American Eel FMP (ASMFC, 2006) and the U.S. Fish and Wildlife Service American Eel Status Review. In addition to management needs, the VIMS Trawl Survey also fulfills data and specimen requests from a variety of agencies, institutes, and individuals for research and educational purposes (Appendix Table 1).

Efforts to improve recruitment indices continue and include evaluation of the size ranges and months (the index period) used in index calculations. A recent VIMS Master's thesis addressed the distributional assumptions of the catch of YOY weakfish and Atlantic croaker (Woodward, 2009). The results showed that weakfish catch data can be described by a gamma distribution and those for Atlantic croaker appear to follow a lognormal distribution. Such findings indicate that indices of abundance calculated for these species could benefit from further refinements. In addition, the use of different index months for weakfish and Atlantic croaker may improve YOY indices by ensuring fewer age 1+ fish are included in YOY index calculations. However, additional work needs to be
conducted to address potential effects of depth on the distribution and catch of these species before refined indices can be recommended and adopted (Woodward, 2009).

The Trawl Survey monitoring program provides more than relative abundance indices used to tune stock assessments and aid in management activities. The data can also be used to investigate factors that influence species abundance that operate on time periods beyond annual recruitment cycles. For example, using fishery-independent survey data from 1968-2004 for estuarinedependent species, Wood and Austin (2009) found that recruitment of anadromous species was negatively correlated with recruitment of species that spawn on the continental shelf. Furthermore, recruitment patterns favored one group over the other for periods greater than a decade and shifts between recruitment regimes occurred within a short period of time (2-3 years; Wood and Austin, 2009). Understanding that long-term recruitment cycles dominate for decades is an important development that affects management options. It may be difficult to observe the effects of regulations aimed at improving recruitment if regulations are enacted during a cycle of low recruitment.

Information from the Trawl Survey also provides a basis for monitoring species interactions. For example, annual catch rates of channel catfish and white catfish have declined since 1991, while catches of the introduced blue catfish have increased dramatically (Connelly, 2001; this report). Because diets and distributions of these species overlap, the observed trends may be due to competition and thus, species interactions warrant further study. Furthermore, the shift in diet of older blue catfish to include other fishes may affect ecosystem function. A recent collaborative summary of blue catfish data from the Chesapeake Bay region was presented at the 2nd International Catfish Symposium and will be published in a conference proceedings journal next year (Schloesser et al. 2010).

Changes in catches of important recreational species may be associated with degradation of estuarine nursery habitats, overfishing, poor recruitment, or a combination of these factors (Murdy et al., 1997). Although it is not possible to determine the cause of recruitment variability from trawl survey data alone, some general observations are possible. Spot recruitment indices have declined greatly over the past 50 years, but year-class strength of this oceanic spawner appears to be controlled by environmental factors occurring outside the Bay (Homer and Mihursky, 1991; Bodolus, 1994). Atlantic croaker recruitment indices show the greatest interannual variability with fluctuations possibly related to environmental conditions that vary annually. Norcross (1983) found that cold winters increased mortality in overwintering juvenile Atlantic croaker and during some years may "push" the spawning population further south, preventing access to nursery areas in Chesapeake Bay. Weakfish indices have remained low since the mid-1990s, and the decline may be attributed to both habitat degradation (loss of seagrass beds in coastal areas) and overfishing. Declines in summer flounder abundance have been observed and may be due to overfishing or year-class failure (Terceiro, 2006). Striped bass display great recruitment variability and one or two strong year classes may dominate the population at any one time (Richards and Rago, 1999). After closure of the fishery in the mid- to late-1980s due to overfishing, poor recruitment, and low stock abundance (Richards and Rago, 1999), the striped bass recruitment index peaked in 1987. Finally, white catfish and channel catfish indices, while variable, have decreased over the past 19 years, possibly due to competition with the introduced blue catfish.

The VIMS trawl survey program supplies critical data for management of fishery resources that use Chesapeake Bay as a spawning or nursery ground. Because the Bay serves as a nursery area for many coastal migratory fish, annual recruitment data are critical for multi-state management efforts along the Atlantic Coast. Furthermore, the trawl survey serves as a foundation to conduct
research on basic biological characteristics of Bay and tributary fishes, as well as a platform from which emerging issues may be addressed.

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TABLES

Table 1. Annual comparisons of substrate (habitat) type from June 1998 - May 2010.

|  | 1998-1999 |  | 1999-2000 |  | 2000-2001 |  | 2001-2002 |  | 2002-2003 |  | 2003-2004 |  | 2004-2005 |  | 2005-2006 |  | 2006-2007 |  | 2007-2008 |  | 2008-2009 |  | 2009-2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Substrate Description | $\begin{aligned} & \text { \%of } \\ & \text { Stations } \end{aligned}$ | Max. Qty. | $\begin{array}{c}\text { \%of } \\ \text { Stations }\end{array}$ | Max. <br> Qty. | $\begin{array}{\|c\|} \hline \% \text { of } \\ \text { Stations } \end{array}$ | Max. Qty. | $\begin{array}{\|c\|} \hline \text { \%of } \\ \text { Stations } \end{array}$ | Max. Qty. | $\begin{array}{\|c\|} \hline \text { \%of } \\ \text { Stations } \end{array}$ | Max. Qty. | \%of Stations | Max. Qty. | \%of Stations | Max. Qty. | \%of Station | Max. <br> Qty. | \%of Stations | Max. <br> Qty. | \%of Stations | Max. Qty. | \%of Stations $^{1}$ | Max. Qty. | \%of Stations $^{1}$ | Max. Qty. |
| Atificial | 0.68 | 2.0 | 3.90 | 3.0 | 4.83 | 2.0 | 5.68 | 15.0 | 5.72 | 4.0 | 7.03 | 7.0 | 3.27 | 2.0 | 1.98 | 2.0 | 2.34 | 3.0 | 4.17 | 16.0 | 4.41 | 12.0 | 2.53 | 12.0 |
| Dead man's fingers (Byyozoan) | 7.21 | 5.0 | 9.54 | 4.0 | 7.42 | 5.0 | 8.89 | 6.0 | 12.16 | 16.0 | 10.29 | 1.0 | 3.76 | 0.5 | 5.04 | 24.0 | 8.30 | 2.0 | 9.97 | 3.0 | 21.40 | 23.0 | 23.28 | 24.0 |
| Detritus | 27.55 | 6.0 | 40.53 | 10.0 | 37.03 | 4.0 | 36.91 | 6.0 | 50.09 | 7.0 | 61.36 | 10.0 | 65.69 | 8.0 | 54.91 | 2.0 | 59.34 | 19.0 | 59.80 | 10.0 | 50.93 | 3.0 | 49.75 | 5.0 |
| Hydroids | 38.59 | 5.0 | 54.19 | 4.0 | 43.20 | 5.0 | 57.01 | 10.5 | 49.36 | 5.0 | 59.15 | 12.0 | 72.63 | 5.0 | 61.77 | 4.0 | 59.09 | 5.0 | 60.29 | 5.0 | 62.21 | 6.0 | 57.68 | 5.0 |
| Sea Squirts (Molgula spp.) | 20.35 | 5.0 | 29.55 | 12.0 | 22.54 | 14.0 | 27.86 | 18.0 | 15.34 | 5.0 | 21.08 | 9.0 | 25.57 | 8.0 | 19.98 | 14.0 | 34.03 | 7.0 | 30.88 | 12.0 | 29.22 | 24.0 | 27.45 | 4.0 |
| Seaweeds | 14.79 | 4.0 | 23.84 | 10.0 | 25.67 | 5.0 | 29.86 | 30.0 | 34.30 | 18.0 | 41.75 | 3.0 | 33.74 | 4.0 | 22.79 | 2.0 | 24.73 | 8.0 | 23.20 | 1.0 | 18.10 | 5.0 | 27.45 | 4.0 |
| Shell (oyster, clam, or mussel) | 18.84 | 3.0 | 26.73 | 4.0 | 22.00 | 5.0 | 29.06 | 8.0 | 32.12 | 4.0 | 25.98 | 3.0 | 21.08 | 4.0 | 23.12 | 12.0 | 21.29 | 2.0 | 21.98 | 8.0 | 21.60 | 4.0 | 20.02 | 3.0 |
| Sponges | 7.96 | 6.0 | 9.47 | 5.0 | 9.84 | 5.0 | 12.33 | 10.0 | 14.34 | 18.0 | 11.03 | 3.0 | 9.48 | 4.0 | 7.10 | 4.0 | 8.21 | 4.0 | 11.11 | 17.0 | 14.54 | 25.0 | 14.46 | 5.0 |
| Submerged Aquatic Vegetation | 4.35 | 3.0 | 8.60 | 1.0 | 10.47 | 2.0 | 5.04 | 2.0 | 6.08 | 1.0 | 3.02 | 0.5 | 6.70 | 2.0 | 2.97 | 0.5 | 4.53 | 1.0 | 5.15 | 0.1 | 4.86 | 0.5 | 7.11 | 1.0 |
| Worm Tubes | 5.33 | 1.0 | 9.54 | 1.0 | 10.47 | 1.0 | 9.29 | 1.0 | 12.34 | 1.0 | 14.05 | 2.0 | 10.87 | 1.0 | 5.45 | 1.0 | 8.05 | 1.0 | 11.11 | 0.5 | 10.83 | 4.0 | 8.82 | 3.0 |
| Mud ${ }^{2}$ | 7.36 | - | 6.50 | - | 7.60 | - | 6.73 | - | 10.89 | - | 8.99 | - | 14.30 | - | 14.37 | - | 6.62 | - | 6.86 | - | 7.22 | - | 5.23 | - |
| Sand ${ }^{2}$ | 10.21 | - | 0.87 | - | 1.43 | - | 1.04 | - | 0.36 | - | 0.49 | - | 0.57 | - | 0.08 | - | 1.00 | - | 8.33 | - | 5.96 | - | 3.59 | - |
| Unknown ${ }^{3}$ | 13.66 | - | 5.42 | - | 4.83 | - | 2.80 | - | 2.72 | - | 1.06 | - | 0.74 | - | 4.38 | - | 6.71 | - | 3.51 | - | 1.85 | - | 3.76 | - |
| NUMBER OF TRAWLS: | 1,332 |  | 1,384 |  | 1,118 |  | 1,249 |  | 1,102 |  | 1,224 |  | 1,224 |  | 1,211 |  | 1,193 |  | 1,224 |  | 1,224 |  | 1,224 |  |

1. Based on the number of occurrences of a habitat type divided by the total number of trawls.
2. Sand and Mud are used when verification can be confimed by direct observation.
3. Unknown is used when none of the categories are found in the traw.

Abundance is estimated relative to the capacity of a cormercial tote (intemal dimensions $25.7^{\prime \prime} \times 16.6^{\prime \prime} \times 10^{\prime \prime}$, approximately 72 liters).
Categores include: $0.5=<1 / 4 \mathrm{bin}, 1=1 / 4 \mathrm{bin}, 2=1 / 2 \mathrm{bin}, 3=3 / 4 \mathrm{bin}, 4=$ full bin, etc.

Table 2. Spatial, temporal, and length (mm) criteria used to calculate recruitment indices.

|  | VIMS Trawl Survey - Area / Time / Size Values by Species |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species - Age | VIMS | Strata Used |  |  |  |  |  |  |  |  | Month |  |  |  |  |  |  |  |  |  |  |  |
|  | SP. CODE | Bay |  |  | James |  | York |  | Rapp |  | Size Cut-off Values (mm) - Darkened Areas Represent Index Months |  |  |  |  |  |  |  |  |  |  |  |
|  |  | B | L | u | L | u | L | u | ᄂ | u | January | February | M arch | April | May | June | July | August | September | October | November | December |
|  |  | - | - | p | - | p | - | p | - | $p$ |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | t | w | p | w | p | w | p | w | p |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | m | e | e | e | e | - | e | - | ' |  |  |  |  |  |  |  |  |  |  |  |  |
| American Eel 1+ | F 0060 |  |  |  |  |  |  |  |  |  | --- | --- | --- | $>152$ | $>152$ | $>152$ | --- | --- | --- | --- | --- | ---- |
| Atlantic Croaker Y-O-Y | 0005 |  |  |  |  |  |  |  |  |  | 0-100 | 0-100 | 0-100 | 0-110 | 0-135 | 0-160 | 0-180 | 0-220 | 0-50 | 0-80 | 0-100 | 0-100 |
| Atlantic Croaker Recruits | F 0005 |  |  |  |  |  |  |  |  |  | 0-100 | 0-100 | 0-100 | 0-110 | 0-135 | 0-160 | 0-180 | 0-220 | 0-50 | 0-80 | 0-100 | 0-100 |
| Bay Anchovy Y-O-Y | 0103 |  |  |  |  |  |  |  |  |  | 0-77 | 0-80 | 0-80 | 0-80 | 0-80 | 0-80 | 0-44 | 0-51 | 0-56 | 0-61 | 0-65 | 0-70 |
| Black Seabass Y-O-Y | F 0002 |  |  |  |  |  |  |  |  |  | 0-110 | 0-110 | 0-110 | 0-110 | 0-110 | 0-150 | 0-175 | 0-70 | 0-85 | 0-100 | 0-105 | 0-110 |
| Blue Catfish Y-O-Y | F 0314 |  |  |  |  |  |  |  |  |  | 0-165 | 0-165 | 0-165 | 0-175 | 0-225 | 0-250 | 0-250 | 0-115 | 0-125 | 0-140 | 0-150 | 0-165 |
| Blue Catfish 1+ | F 0314 |  |  |  |  |  |  |  |  |  | $>165$ | $>165$ | >165 | >175 | >225 | >250 | >250 | >115 | $>125$ | >140 | >150 | $>165$ |
| Channel Catfish Y-O-Y | F 0040 |  |  |  |  |  |  |  |  |  | 0-130 | 0-130 | 0-130 | 0-140 | 0-150 | 0-50 | 0-80 | 0-105 | 0-120 | 0-130 | 0-130 | 0-130 |
| Channel Catfish 1+ | F 0040 |  |  |  |  |  |  |  |  |  | $>130$ | >130 | >130 | >140 | >150 | $>50$ | >80 | >105 | $>120$ | >130 | >130 | >130 |
| Scup | F 0001 |  |  |  |  |  |  |  |  |  | 90-170 | 90-170 | 90-170 | 90-170 | 35-90 | 40-100 | 50-125 | 60-145 | 75-160 | 85-170 | 90-170 | 90-170 |
| Silver Perch Y-O-Y | F 0213 |  |  |  |  |  |  |  |  |  | 0-160 | 0-160 | 0-160 | 0-160 | 0-165 | 0-170 | 0-100 | 0-130 | 0-150 | 0-160 | 0-160 | 0-160 |
| Spot Y-O-Y | 0033 |  |  |  |  |  |  |  |  |  | 0-200 | 0-200 | 0-50 | 0-75 | 0-100 | 0-135 | 0-160 | 0-180 | 0-200 | 0-200 | 0-200 | 0-200 |
| Striped Bass Y-O-Y | F 0031 |  |  |  |  |  |  |  |  |  | 0-200 | 0-200 | 0-200 | 0-200 | 0-50 | 0-80 | 0-100 | 0-120 | 0-135 | 0-150 | 0-175 | 0-190 |
| Summer Flounder Y-O-Y | F 0003 |  |  |  |  |  |  |  |  |  | 0-290 | 0-290 | 0-60 | 0-100 | 0-140 | 0-170 | 0-200 | 0-225 | 0-250 | 0-275 | 0-290 | 0-290 |
| Weakfish Y-O-Y | F 0007 |  |  |  |  |  |  |  |  |  | 0-200 | 0-200 | 0-200 | 0-225 | 0-240 | 0-90 | 0-120 | 0-150 | 0-180 | 0-200 | 0-200 | 0-200 |
| White Catfish Y-O-Y | F 0039 |  |  |  |  |  |  |  |  |  | 0-110 | 0-110 | 0-110 | 0-110 | 0-120 | 0-50 | 0-65 | 0-80 | 0-90 | 0-100 | 0-110 | 0-110 |
| White Catfish 1+ | F 0039 |  |  |  |  |  |  |  |  |  | $>110$ | $>110$ | $>110$ | $>110$ | $>120$ | >50 | >65 | >80 | >90 | >100 | >110 | $>110$ |
| White Perch Y-O-Y | F 0032 |  |  |  |  |  |  |  |  |  | 0-85 | 0-85 | 0-85 | 0-95 | 0-35 | 0-65 | 0-73 | 0-80 | 0-85 | 0-85 | 0-85 | 0-85 |
| White Perch 1+ | F 0032 |  |  |  |  |  |  |  |  |  | 86-300 | 86-300 | 86-300 | 96-300 | 36-300 | 66-300 | 74-300 | 81-300 | 86-300 | 85-300 | 86-300 | 86-300 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3. Sample collection history of the VIMS Trawl Survey, 1988 - May 2010. Each entry in the table represents the number of completed tows; YR is year, TOT is total, STAT. TYPE is station type. Other codes are below and are based on Wojcik and Van Engel (1988). Anvendices A - C

| YR | TOT | MONTH |  |  |  |  |  |  |  |  |  |  |  | WATER SYSTEM |  |  |  |  |  | Vessel |  |  | Gear |  | STAT. TYPE |  | TOW DURATION/DISTANCE |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | J | F | M | A | M | J | J | A | S | 0 | N | D | CL | JA | PO | RA | YK | ZZ | FH | JS | LN | 070 | 108 | F | R | 5 | OT | DIS |
| 1988 | 889 | 69 | 69 | 62 | 48 | 82 | 82 | 82 | 82 | 82 | 82 | 80 | 69 | 576 | 97 | 0 | 105 | 111 | 0 | 0 | 889 | 0 | 889 | 0 | 313 | 576 | 885 | 0 | 4 |
| 1989 | 840 | 61 | 61 | 61 | 66 | 76 | 76 | 76 | 76 | 76 | 76 | 76 | 59 | 479 | 108 | 0 | 124 | 129 | 0 | 0 | 840 | 0 | 840 | 0 | 361 | 479 | 840 | 0 | 0 |
| 1990 | 827 | 61 | 61 | 61 | 61 | 76 | 76 | 77 | 75 | 76 | 69 | 76 | 58 | 473 | 108 | 0 | 119 | 127 | 0 | 279 | 548 | 0 | 827 | 0 | 354 | 473 | 826 | 0 | 1 |
| 1991 | 930 | 61 | 25 | 61 | 61 | 73 | 94 | 95 | 95 | 97 | 97 | 97 | 74 | 411 | 108 | 0 | 120 | 291 | 0 | 930 | 0 | 0 | 0 | 930 | 357 | 573 | 928 | 1 | 1 |
| 1992 | 982 | 79 | 47 | 79 | 79 | 97 | 88 | 88 | 88 | 89 | 88 | 88 | 72 | 404 | 110 | 0 | 124 | 344 | 0 | 982 | 0 | 0 | 0 | 982 | 361 | 621 | 975 | 7 | 0 |
| 1993 | 915 | 40 | 73 | 40 | 71 | 88 | 89 | 88 | 88 | 88 | 88 | 87 | 75 | 370 | 110 | 0 | 126 | 309 | 0 | 915 | 0 | 0 | 0 | 915 | 365 | 550 | 914 | 1 | 0 |
| 1994 | 911 | 40 | 73 | 40 | 73 | 88 | 88 | 88 | 88 | 88 | 88 | 88 | 69 | 368 | 110 | 0 | 124 | 309 | 0 | 911 | 0 | 0 | 0 | 911 | 363 | 548 | 906 | 5 | 0 |
| 1995 | 993 | 40 | 73 | 40 | 73 | 92 | 88 | 88 | 88 | 105 | 105 | 99 | 102 | 411 | 96 | 0 | 201 | 285 | 0 | 993 | 0 | 0 | 0 | 993 | 314 | 679 | 984 | 9 | 0 |
| 1996 | 1176 | 52 | 91 | 71 | 106 | 106 | 107 | 108 | 108 | 107 | 108 | 107 | 105 | 435 | 228 | 0 | 258 | 255 | 0 | 1176 | 0 | 0 | 0 | 1176 | 279 | 897 | 1168 | 6 | 2 |
| 1997 | 1220 | 68 | 105 | 66 | 98 | 110 | 111 | 111 | 112 | 111 | 112 | 111 | 105 | 425 | 265 | 0 | 264 | 266 | 0 | 1220 | 0 | 0 | 0 | 1220 | 302 | 918 | 1217 | 3 | 0 |
| 1998 | 1262 | 66 | 105 | 66 | 105 | 111 | 111 | 128 | 59 | 138 | 124 | 130 | 119 | 388 | 265 | 0 | 256 | 264 | 89 | 1262 | 0 | 0 | 0 | 1262 | 322 | 940 | 1261 | 1 | 0 |
| 1999 | 1382 | 79 | 122 | 80 | 122 | 120 | 118 | 119 | 118 | 122 | 124 | 131 | 127 | 402 | 264 | 0 | 264 | 265 | 187 | 1382 | 0 | 0 | 0 | 1382 | 363 | 1019 | 1380 | 2 | 0 |
| 2000 | 1367 | 52 | 129 | 85 | 101 | 158 | 111 | 128 | 125 | 121 | 141 | 111 | 105 | 433 | 250 | 17 | 266 | 265 | 136 | 1367 | 0 | 0 | 0 | 1367 | 363 | 1004 | 1367 | 0 | 0 |
| 2001 | 1122 | 30 | 30 | 30 | 75 | 112 | 144 | 111 | 112 | 135 | 136 | 111 | 96 | 384 | 230 | 35 | 230 | 230 | 13 | 1017 | 0 | 105 | 0 | 1122 | 277 | 845 | 1119 | 1 | 2 |
| 2002 | 1090 | 66 | 90 | 66 | 90 | 96 | 106 | 96 | 97 | 95 | 96 | 96 | 96 | 288 | 264 | 0 | 264 | 264 | 10 | 1090 | 0 | 0 | 0 | 1090 | 300 | 790 | 1089 | 1 | 0 |
| 2003 | 1191 | 66 | 96 | 66 | 96 | 96 | 111 | 111 | 111 | 111 | 111 | 111 | 105 | 399 | 264 | 0 | 264 | 264 | 0 | 1191 | 0 | 0 | 0 | 1191 | 300 | 891 | 1191 | 0 | 0 |
| 2004 | 1224 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 111 | 111 | 111 | 105 | 432 | 264 | 0 | 264 | 264 | 0 | 1224 | 0 | 0 | 0 | 1224 | 300 | 924 | 1224 | 0 | 0 |
| 2005 | 1211 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 113 | 111 | 111 | 90 | 419 | 264 | 0 | 264 | 264 | 0 | 1211 | 0 | 0 | 0 | 1211 | 300 | 911 | 1211 | 0 | 0 |
| 2006 | 1193 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 113 | 111 | 78 | 105 | 423 | 242 | 0 | 264 | 264 | 0 | 1193 | 0 | 0 | 0 | 1193 | 292 | 901 | 1193 | 0 | 0 |
| 2007 | 1224 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 111 | 111 | 111 | 105 | 432 | 264 | 0 | 264 | 264 | 0 | 1224 | 0 | 0 | 0 | 1224 | 300 | 924 | 1224 | 0 | 0 |
| 2008 | 1224 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 111 | 111 | 111 | 105 | 432 | 264 | 0 | 264 | 264 | 0 | 1224 | 0 | 0 | 0 | 1224 | 300 | 924 | 1224 | 0 | 0 |
| 2009 | 1224 | 66 | 105 | 66 | 105 | 111 | 111 | 111 | 111 | 111 | 111 | 111 | 105 | 432 | 264 | 0 | 264 | 264 | 0 | 1224 | 0 | 0 | 0 | 1224 | 300 | 924 | 1224 | 0 | 0 |
| 2010 | 504 | 66 | 105 | 83 | 122 | 138 |  |  |  |  |  |  |  | 123 | 110 | 0 | 110 | 110 | 51 | 504 | 0 | 0 | 0 | 504 | 146 | 358 | 504 | 0 | 0 |
| TOT | 24901 | 1392 | 1985 | 1453 | 2077 | 2385 | 2266 | 2260 | 2188 | 2311 | 2311 | 2232 | 2051 | 9339 | 4549 |  | 4803 | 5672 | 486 | 22519 | 2277 | 105 | 2556 ${ }^{\prime \prime}$ | 22345 | 7232 | 17669 | 24854 | 37 | 10 |

System:

| CL | Lower Chesapeake Bay (Virginia Portion) |
| :--- | :--- |
| JA | James River |
| PO | Potomac River |
| RA | Rappahannock River |
| YK | York River |
| ZZ | includes: Atlantic Ocean (AT) - 1971, 78-79, 2002; Piankatank R. (PK) - 1970-71, 98-00; |
|  | Mobjack Bay (MB) - 1970-73, 98-01, 2010; Pocomoke Sound (CP) -1973-81, 98-01; |
|  | Great Wicomico R. (GW) -1998-00. |


| Vessel: | FH | Fish Hawk |
| :---: | :--- | :--- |
| JS | John Smith |  |
| LN | Langley II |  |

Vessel: FH Fish Hawk
LN Langley II
RA Rappahannock River
ZZ includes: Atlantic Ocean (AT) - 1971, 78-79, 2002; Piankatank R. (PK) - 1970-71, 98-00 Great Wicomico R. (GW) - 1998-00.

| Gear Code: |  |  | Station Type: |
| :--- | :--- | :--- | :--- |
| 30' Gears | 070 | Lined, tickler chain, $60^{\prime}$ bridle, $54 " x 24 "$ doors | R - Random |
|  | 108 | Lined, tickler chain, $60^{\prime}$ bridle, metal china-v doors |  |

Tow Type: OT is tow duration in minutes for those not listed.
DIS is distance.

Table 4. VIMS trawl survey pooled catch for June 2009 to May 2010 from 1,224 tows. Adjusted Percent of Catch Excludes Bay Anchow and Hogchoker

| Species | Number of Fish (All) | Frequency | Percent of Catch | Catch Per <br> Trawl | Adjusted Percent of Catch | Number of Fish YOY | Average Length (mm) | Standard Error (length) | $\begin{gathered} \hline \text { Minimum } \\ \text { Length } \\ (\mathrm{mm}) \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Maximum } \\ \text { Length } \\ (\mathrm{mm}) \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| bay anchow | 338,113 | 1056 | 60.09 | 276.24 |  | 253,549 | 55 | 0.06 | 21 | 103 |
| hogchoker | 79,161 | 747 | 14.07 | 64.67 |  | 18,341 | 83 | 0.17 | 22 | 193 |
| Atlantic croaker | 57,660 | 797 | 10.25 | 47.11 | 39.67 | 45,206 | 110 | 0.43 | 13 | 466 |
| white perch | 16,547 | 284 | 2.94 | 13.52 | 11.38 | 4,354 | 129 | 0.56 | 19 | 286 |
| spot | 15,341 | 644 | 2.73 | 12.53 | 10.55 | 12,099 | 131 | 0.35 | 14 | 237 |
| weakfish | 14,048 | 554 | 2.50 | 11.48 | 9.66 | 12,681 | 107 | 0.42 | 19 | 319 |
| blue catish | 9,526 | 218 | 1.69 | 7.78 | 6.55 | 2,722 | 230 | 1.01 | 27 | 895 |
| blackcheek tonguefish | 4,297 | 402 | 0.76 | 3.51 | 2.96 | 532 | 130 | 0.46 | 24 | 193 |
| northern searobin | 3,158 | 178 | 0.56 | 2.58 | 2.17 | 3,133 | 100 | 0.39 | 21 | 189 |
| striped anchovy | 2,737 | 152 | 0.49 | 2.24 | 1.88 | 2,714 | 83 | 0.47 | 34 | 132 |
| spotted hake | 2,618 | 241 | 0.47 | 2.14 | 1.80 | 2,584 | 128 | 0.81 | 38 | 284 |
| silver perch | 2,352 | 268 | 0.42 | 1.92 | 1.62 | 2,114 | 123 | 0.59 | 24 | 199 |
| kingfish spp | 1,759 | 238 | 0.31 | 1.44 | 1.21 | 1,707 | 75 | 0.78 | 23 | 277 |
| scup | 1,678 | 104 | 0.30 | 1.37 | 1.15 | 1,663 | 93 | 0.33 | 49 | 152 |
| gizzard shad | 1,474 | 192 | 0.26 | 1.2 | 1.01 | 613 | 208 | 1.96 | 83 | 405 |
| Atlantic menhaden | 1,299 | 236 | 0.23 | 1.06 | 0.89 | 1,096 | 85 | 1.86 | 27 | 320 |
| smallmouth flounder | 1,097 | 175 | 0.19 | 0.9 | 0.75 | 864 | 83 | 0.5 | 38 | 139 |
| summer flounder | 1,095 | 406 | 0.19 | 0.89 | 0.75 | 685 | 239 | 2.92 | 25 | 670 |
| southern kingfish | 937 | 179 | 0.17 | 0.77 | 0.64 | 803 | 120 | 1.69 | 35 | 303 |
| blueback herring | 923 | 129 | 0.16 | 0.75 | 0.63 | 921 | 76 | 0.41 | 29 | 177 |
| harvestfish | 810 | 178 | 0.14 | 0.66 | 0.56 | 788 | 64 | 0.99 | 13 | 172 |
| striped bass | 676 | 122 | 0.12 | 0.55 | 0.47 | 568 | 138 | 3.54 | 19 | 623 |
| threadfin shad | 598 | 37 | 0.11 | 0.49 | 0.41 |  | 88 | 0.67 | 58 | 119 |
| butterfish | 584 | 134 | 0.10 | 0.48 | 0.40 | 493 | 86 | 1.45 | 14 | 192 |
| oyster toadfish | 541 | 153 | 0.10 | 0.44 | 0.37 |  | 168 | 3.18 | 24 | 377 |
| alewife | 392 | 75 | 0.07 | 0.32 | 0.27 | 383 | 117 | 1.02 | 74 | 255 |
| white catfish | 368 | 106 | 0.07 | 0.3 | 0.25 | 153 | 175 | 5.04 | 56 | 428 |
| Atlantic silverside | 345 | 65 | 0.06 | 0.28 | 0.24 | 345 | 92 | 0.62 | 51 | 119 |
| northern pipefish | 310 | 157 | 0.06 | 0.25 | 0.21 |  | 142 | 1.99 | 52 | 245 |
| black seabass | 180 | 95 | 0.03 | 0.15 | 0.12 | 127 | 108 | 2.76 | 31 | 221 |
| seaboard goby | 167 | 39 | 0.03 | 0.14 | 0.11 |  | 37 | 0.71 | 22 | 60 |
| striped searobin | 156 | 79 | 0.03 | 0.13 | 0.11 |  | 73 | 2.76 | 29 | 188 |
| naked goby | 133 | 68 | 0.02 | 0.11 | 0.09 |  | 40 | 0.65 | 19 | 56 |
| American eel | 124 | 73 | 0.02 | 0.1 | 0.09 |  | 298 | 9.66 | 53 | 658 |
| northern puffer | 103 | 71 | 0.02 | 0.08 | 0.07 | 72 | 110 | 3.86 | 16 | 205 |
| clearnose skate | 103 | 46 | 0.02 | 0.08 | 0.07 |  | 430 | 5.63 | 105 | 528 |
| lined seahorse | 98 | 71 | 0.02 | 0.08 | 0.07 |  | 77 | 2.37 | 38 | 126 |
| red hake | 93 | 23 | 0.02 | 0.08 | 0.06 |  | 138 | 2.98 | 65 | 200 |
| banded drum | 86 | 28 | 0.02 | 0.07 | 0.06 |  | 78 | 5.97 | 22 | 229 |
| Atlantic thread herring | 72 | 27 | 0.01 | 0.06 | 0.05 |  | 89 | 4.61 | 40 | 182 |
| inshore lizardfish | 68 | 42 | 0.01 | 0.06 | 0.05 | 53 | 151 | 7.39 | 53 | 314 |
| windowpane | 57 | 35 | 0.01 | 0.05 | 0.04 | 49 | 141 | 7.84 | 48 | 291 |
| American shad | 57 | 32 | 0.01 | 0.05 | 0.04 | 56 | 116 | 1.87 | 82 | 150 |
| skilletfish | 55 | 31 | 0.01 | 0.04 | 0.04 |  | 39 | 1.95 | 19 | 67 |
| Atlantic moonfish | 55 | 23 | 0.01 | 0.04 | 0.04 |  | 79 | 2.23 | 44 | 126 |
| bluefish | 50 | 31 | 0.01 | 0.04 | 0.03 |  | 168 | 6.03 | 119 | 276 |
| Atlantic cutlassfish | 46 | 24 | 0.01 | 0.04 | 0.03 |  | 347 | 18.83 | 106 | 607 |
| feather blenny | 40 | 27 | 0.01 | 0.03 | 0.03 |  | 69 | 3.08 | 31 | 109 |
| striped mullet | 31 | 12 | 0.01 | 0.03 | 0.02 |  | 205 | 7.32 | 151 | 326 |
| pigfish | 30 | 18 | 0.01 | 0.02 | 0.02 |  | 146 | 4.53 | 106 | 190 |
| sea lamprey | 28 | 21 | 0.00 | 0.02 | 0.02 |  | 162 | 2.84 | 137 | 201 |
| Atlantic spadefish | 25 | 20 | 0.00 | 0.02 | 0.02 |  | 75 | 6.69 | 24 | 155 |
| silver hake | 24 | 12 | 0.00 | 0.02 | 0.02 |  | 150 | 5.99 | 90 | 192 |
| striped cusk-eel | 23 | 6 | 0.00 | 0.02 | 0.02 |  | 105 | 5.89 | 84 | 211 |
| red drum | 19 | 13 | 0.00 | 0.02 | 0.01 |  | 67 | 4.42 | 47 | 113 |
| channel catfish | 19 | 11 | 0.00 | 0.02 | 0.01 | 1 | 303 | 16.88 | 80 | 505 |
| hickory shad | 18 | 14 | 0.00 | 0.01 | 0.01 |  | 128 | 16.06 | 84 | 318 |
| spotted seatrout | 18 | 14 | 0.00 | 0.01 | 0.01 |  | 176 | 8.92 | 128 | 256 |
| longnose gar | 17 | 15 | 0.00 | 0.01 | 0.01 |  | 743 | 49.2 | 353 | 1045 |
| Spanish mackerel | 13 | 9 | 0.00 | 0.01 | 0.01 |  | 116 | 16.82 | 37 | 219 |
| green goby | 11 | 10 | 0.00 | 0.01 | 0.01 |  | 43 | 1.62 | 35 | 52 |
| bluntnose stingray | 11 | 6 | 0.00 | 0.01 | 0.01 |  | 426 | 54.18 | 196 | 780 |
| common carp | 10 | 8 | 0.00 | 0.01 | 0.01 | . | 490 | 40.93 | 192 | 660 |

Table 4 (continued)
Adjusted Percent of Catch Excludes Bay Anchowy and Hogchoker

| Species | Number of Fish (All) | Frequency | Percent of Catch | Catch Per <br> Trawl | Adjusted Percent of Catch | Number of Fish YOY | Average Length (mm) | Standard Error (length) | Minimum <br> Length <br> (mm) | Maximum <br> Length (mm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| spiny butterfly ray | 10 | 7 | 0 | 0.01 | 0.01 |  | 591 | 81.26 | 473 | 1320 |
| black drum | 9 | 8 | 0 | 0.01 | 0.01 |  | 207 | 10.12 | 151 | 249 |
| northern stargazer | 9 | 8 | 0 | 0.01 | 0.01 |  | 56 | 22.36 | 23 | 234 |
| northern sand lance | 8 | 3 | 0 | 0.01 | 0.01 |  | 140 | 7.18 | 115 | 166 |
| Atlantic stingray | 7 | 7 | 0 | 0.01 | 0 |  | 278 | 36.1 | 172 | 401 |
| northern kingfish | 7 | 6 | 0 | 0.01 | 0 |  | 105 | 17.55 | 45 | 164 |
| smooth dogfish | 6 | 6 | 0 | 0 | 0 |  | 456 | 38.57 | 343 | 580 |
| smooth butterfly ray | 6 | 6 | 0 | 0 | 0 |  | 436 | 47.26 | 270 | 630 |
| spottail shiner | 6 | 4 | 0 | 0 | 0 |  | 87 | 2.3 | 78 | 95 |
| rough silverside | 6 | 4 | 0 | 0 | 0 |  | 62 | 6.62 | 44 | 92 |
| striped burrish | 5 | 4 | 0 | 0 | 0 |  | 140 | 34.12 | 31 | 232 |
| Atlantic herring | 4 | 4 | 0 | 0 | 0 |  | 49 | 3.99 | 40 | 59 |
| brown bullhead | 4 | 4 | 0 | 0 | 0 |  | 148 | 33.48 | 81 | 227 |
| conger eel | 4 | 4 | 0 | 0 | 0 |  | 519 | 57.44 | 447 | 690 |
| lookdown | 4 | 4 | 0 | 0 | 0 |  | 99 | 8.44 | 86 | 124 |
| silver seatrout | 4 | 4 | 0 | 0 | 0 |  | 148 | 15.12 | 123 | 191 |
| tessellated darter | 4 | 3 | 0 | 0 | 0 |  | 72 | 2.78 | 66 | 79 |
| blue runner | 4 | 3 | 0 | 0 | 0 |  | 149 | 7.31 | 128 | 159 |
| dusky pipefish | 3 | 3 | 0 | 0 | 0 |  | 101 | 3.06 | 97 | 107 |
| star drum | 3 | 3 | 0 | 0 | 0 |  | 81 | 1.86 | 79 | 85 |
| inland silverside | 3 | 2 | 0 | 0 | 0 |  | 66 | 1.76 | 63 | 69 |
| Atlantic bumper | 3 | 2 | 0 | 0 | 0 |  | 69 | 2 | 65 | 71 |
| gulf kingfish | 3 | 1 | 0 | 0 | 0 |  | 127 | 13.37 | 101 | 146 |
| round herring | 2 | 2 | 0 | 0 | 0 |  | 235 | 42.5 | 192 | 277 |
| eastern silvery minnow | 2 | 2 | 0 | 0 | 0 |  | 91 | 5.5 | 85 | 96 |
| banded killifish | 2 | 2 | 0 | 0 | 0 |  | 44 | 3 | 41 | 47 |
| striped killifish | 2 | 2 | 0 | 0 | 0 |  | 44 | 8.5 | 35 | 52 |
| bluegill | 2 | 2 | 0 | 0 | 0 |  | 43 | 3.5 | 39 | 46 |
| roughtail stingray | 2 | 2 | 0 | 0 | 0 |  | 447 | 60 | 387 | 507 |
| rough scad | 2 | 2 | 0 | 0 | 0 |  | 57 | 4.5 | 52 | 61 |
| fringed flounder | 2 | 2 | 0 | 0 | 0 |  | 95 | 6 | 89 | 101 |
| chain pipefish | 2 | 2 | 0 | 0 | 0 |  | 224 | 48.5 | 175 | 272 |
| striped blenny | 2 | 1 | 0 | 0 | 0 |  | 30 | 0 | 30 | 30 |
| winter flounder | 1 | 1 | 0 | 0 | 0 |  | 86 | . | 86 | 86 |
| tautog | 1 | 1 | 0 | 0 | 0 |  | 136 | . | 136 | 136 |
| bluespotted cornetfish | 1 | 1 | 0 | 0 | 0 |  | 284 | . | 284 | 284 |
| threespine stickleback | 1 | 1 | 0 | 0 | 0 |  | 67 | . | 67 | 67 |
| pumpkinseed | 1 | 1 | 0 | 0 | 0 |  | 58 | . | 58 | 58 |
| pinfish | 1 | 1 | 0 | 0 | 0 |  | 131 | . | 131 | 131 |
| planehead filefish | 1 | 1 | 0 | 0 | 0 |  | 40 |  | 40 | 40 |
| bluespotted sunfish | 1 | 1 | 0 | 0 | 0 |  | 46 | . | 46 | 46 |
| spotfin mojarra | 1 | 1 | 0 | 0 | 0 |  | 61 | . | 61 | 61 |

Table 5. American eel indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. <br> Mean | 95\% C.I.'s | C.V. | N | River Only (RO) | N |
| 1988 | 1.26 | 0.48-2.46 | 26.08 | 18 | 2.30 | 18 |
| 1989 | 7.93 | 4.62-13.18 | 10.57 | 31 | 8.82 | 31 |
| 1990 | 4.85 | 3.25-7.04 | 9.02 | 30 | 6.67 | 31 |
| 1991 | 2.06 | 0.70-4.50 | 26.34 | 31 | 2.12 | 31 |
| 1992 | 8.29 | 6.24-10.92 | 5.60 | 31 | 4.01 | 31 |
| 1993 | 3.56 | 2.44-5.05 | 9.29 | 31 | 3.68 | 31 |
| 1994 | 2.23 | 1.06-4.08 | 19.25 | 31 | 2.48 | 31 |
| 1995 | 2.69 | 2.01-3.53 | 7.83 | 33 | 2.44 | 33 |
| 1996 | 2.57 | 1.77-3.59 | 9.94 | 84 | 2.81 | 33 |
| 1997 | 2.29 | 1.11-4.13 | 18.69 | 90 | 1.37 | 39 |
| 1998 | 2.00 | 1.00-3.51 | 18.49 | 90 | 2.30 | 39 |
| 1999 | 1.25 | 0.58-2.19 | 21.67 | 90 | 1.14 | 39 |
| 2000 | 1.42 | 0.75-2.35 | 18.42 | 90 | 1.15 | 38 |
| 2001 | 0.79 | 0.18-1.72 | 35.92 | 90 | 0.46 | 39 |
| 2002 | 0.80 | 0.30-1.52 | 28.11 | 90 | 0.93 | 39 |
| 2003 | 0.79 | 0.22-1.61 | 32.68 | 90 | 0.60 | 39 |
| 2004 | 0.43 | 0.21-0.68 | 22.95 | 90 | 0.50 | 39 |
| 2005 | 0.35 | 0.21-0.51 | 18.66 | 90 | 0.47 | 39 |
| 2006 | 0.15 | 0.00-0.32 | 49.69 | 90 | 0.06 | 39 |
| 2007 | 0.22 | 0.07-0.38 | 31.92 | 90 | 0.14 | 39 |
| 2008 | 0.25 | 0.04-0.49 | 40.09 | 90 | 0.23 | 39 |
| 2009 | 0.57 | 0.23-1.01 | 27.37 | 90 | 0.34 | 39 |

Table 6. Spring Atlantic croaker indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | Bay \& Rive <br> (BRI) | River Only |  |  |
|  | Mean | 95\% C.I.'s |  |  |  | N | (RO) | N |
| 1988 | 0.36 | 0.21-0.44 | 16.05 | 234 | 0.38 | 234 | 2.22 | 84 |
| 1989 | 0.65 | 0.38-0.85 | 15.63 | 252 | 0.78 | 252 | 4.63 | 84 |
| 1990 | 0.48 | 0.23-0.67 | 20.56 | 252 | 0.52 | 252 | 2.98 | 85 |
| 1991 | 4.12 | 2.83-5.84 | 8.87 | 307 | 4.35 | 238 | 12.87 | 83 |
| 1992 | 1.17 | 0.77-1.67 | 13.17 | 309 | 1.34 | 240 | 10.26 | 84 |
| 1993 | 1.98 | 1.33-2.80 | 11.20 | 301 | 2.21 | 240 | 19.40 | 84 |
| 1994 | 0.86 | 0.56-1.22 | 14.33 | 300 | 0.95 | 240 | 2.98 | 84 |
| 1995 | 0.95 | 0.67-1.28 | 11.55 | 306 | 0.93 | 246 | 5.55 | 90 |
| 1996 | 0.19 | 0.11-0.28 | 19.63 | 405 | 0.16 | 242 | 0.36 | 88 |
| 1997 | 1.47 | 1.15-1.85 | 7.78 | 419 | 0.87 | 255 | 7.78 | 100 |
| 1998 | 1.19 | 0.95-1.47 | 7.51 | 374 | 0.48 | 214 | 6.21 | 96 |
| 1999 | 1.50 | 1.05-2.05 | 10.83 | 397 | 1.28 | 232 | 4.08 | 100 |
| 2000 | 0.60 | 0.42-0.80 | 12.68 | 413 | 0.44 | 245 | 1.39 | 97 |
| 2001 | 0.37 | 0.25-0.49 | 14.38 | 420 | 0.32 | 256 | 1.18 | 100 |
| 2002 | 1.59 | 1.07-2.22 | 11.59 | 361 | 1.11 | 197 | 4.80 | 100 |
| 2003 | 0.49 | 0.28-0.74 | 19.19 | 405 | 0.52 | 241 | 0.28 | 100 |
| 2004 | 0.96 | 0.73-1.22 | 9.34 | 420 | 0.70 | 255 | 4.42 | 99 |
| 2005 | 0.47 | 0.35-0.59 | 10.46 | 420 | 0.31 | 256 | 1.85 | 100 |
| 2006 | 1.27 | 1.00-1.59 | 7.90 | 420 | 0.77 | 256 | 3.92 | 100 |
| 2007 | 1.04 | 0.76-1.37 | 10.34 | 420 | 0.76 | 256 | 3.05 | 100 |
| 2008 | 4.10 | 3.04-5.45 | 7.16 | 420 | 2.84 | 256 | 18.62 | 100 |
| 2009 | 1.82 | 1.41-2.30 | 7.65 | 420 | 1.48 | 256 | 10.06 | 100 |

Table 7. Bay anchovy indices (1988-2009).

|  | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | Bay \& River |  | ver O |  |
| Year | Mean | 95\% C.I.'s | C.V. | N | (BRI) | N | (RO) | N |
| 1988 | 18.25 | 12.17-27.15 | 6.42 | 346 | 18.06 | 346 | 32.66 | 128 |
| 1989 | 52.47 | 36.27-75.71 | 4.54 | 374 | 51.59 | 374 | 22.74 | 128 |
| 1990 | 6.79 | 4.41-10.22 | 8.89 | 369 | 6.65 | 369 | 8.78 | 124 |
| 1991 | 22.51 | 15.05-33.43 | 6.04 | 491 | 22.83 | 350 | 33.41 | 125 |
| 1992 | 40.14 | 27.17-59.09 | 5.10 | 448 | 40.79 | 355 | 14.53 | 128 |
| 1993 | 43.31 | 28.80-64.89 | 5.23 | 449 | 42.71 | 360 | 28.93 | 132 |
| 1994 | 14.67 | 9.93-21.46 | 6.54 | 444 | 14.36 | 354 | 19.86 | 130 |
| 1995 | 18.36 | 12.84-26.07 | 5.66 | 540 | 18.52 | 362 | 18.57 | 138 |
| 1996 | 15.31 | 11.20-20.82 | 5.21 | 607 | 16.91 | 363 | 5.11 | 135 |
| 1997 | 18.96 | 13.63-26.23 | 5.19 | 625 | 17.33 | 378 | 12.64 | 150 |
| 1998 | 30.26 | 20.75-43.93 | 5.27 | 579 | 30.47 | 336 | 9.70 | 146 |
| 1999 | 15.47 | 11.20-21.22 | 5.35 | 606 | 14.38 | 360 | 21.26 | 150 |
| 2000 | 36.58 | 26.69-49.99 | 4.21 | 619 | 40.36 | 369 | 16.24 | 147 |
| 2001 | 9.55 | 6.93-13.04 | 6.06 | 627 | 9.23 | 377 | 4.56 | 150 |
| 2002 | 5.51 | 3.58-8.24 | 9.36 | 540 | 4.09 | 294 | 9.30 | 150 |
| 2003 | 18.03 | 13.17-24.56 | 5.01 | 624 | 20.65 | 378 | 3.41 | 150 |
| 2004 | 23.06 | 16.71-31.70 | 4.82 | 624 | 21.45 | 377 | 7.02 | 149 |
| 2005 | 22.27 | 16.01-30.85 | 4.98 | 613 | 21.26 | 367 | 8.43 | 150 |
| 2006 | 19.31 | 14.00-26.50 | 5.03 | 592 | 16.99 | 360 | 10.59 | 142 |
| 2007 | 23.76 | 17.33-32.44 | 4.69 | 624 | 21.15 | 378 | 10.27 | 150 |
| 2008 | 50.29 | 36.21-69.68 | 4.07 | 624 | 43.11 | 378 | 49.06 | 150 |
| 2009 | 30.12 | 22.30-40.55 | 4.21 | 624 | 25.64 | 378 | 25.09 | 150 |

Table 8. Black sea bass indices (1988-2008).


Table 9. Blue catfish juvenile indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | River Onl |  |
|  | Mean | 95\% C.I.'s | C.V. | N | (RO) | N |
| 1988 | 0.00 | 0.00-0.01 | 100.00 | 52 | 0.01 | 52 |
| 1989 | 0.43 | 0.00-1.24 | 61.91 | 51 | 0.25 | 52 |
| 1990 | 0.14 | 0.02-0.28 | 42.14 | 52 | 0.29 | 52 |
| 1991 | 0.39 | 0.27-0.53 | 14.11 | 52 | 0.19 | 52 |
| 1992 | 0.36 | 0.17-0.59 | 24.87 | 52 | 0.26 | 52 |
| 1993 | 0.18 | 0.07-0.30 | 28.51 | 52 | 0.45 | 52 |
| 1994 | 0.17 | 0.03-0.33 | 40.81 | 52 | 0.38 | 52 |
| 1995 | 0.64 | 0.34-1.00 | 20.18 | 109 | 0.91 | 52 |
| 1996 | 0.92 | 0.40-1.63 | 24.21 | 120 | 1.24 | 52 |
| 1997 | 2.40 | 1.55-3.54 | 11.81 | 120 | 2.33 | 52 |
| 1998 | 0.31 | 0.14-0.52 | 26.57 | 120 | 0.54 | 52 |
| 1999 | 0.14 | 0.04-0.25 | 36.47 | 114 | 0.30 | 52 |
| 2000 | 0.22 | 0.00-0.60 | 66.93 | 120 | 0.10 | 52 |
| 2001 | 0.02 | 0.00-0.04 | 67.15 | 120 | 0.02 | 52 |
| 2002 | 0.61 | 0.14-1.28 | 36.48 | 120 | 0.50 | 52 |
| 2003 | 1.33 | 0.75-2.10 | 16.88 | 120 | 2.50 | 52 |
| 2004 | 1.82 | 0.83-3.35 | 20.95 | 120 | 3.34 | 52 |
| 2005 | 2.59 | 1.53-4.10 | 13.75 | 120 | 3.88 | 52 |
| 2006 | 1.52 | 0.74-2.65 | 19.95 | 120 | 3.34 | 52 |
| 2007 | 0.60 | 0.18-1.17 | 32.32 | 120 | 0.66 | 52 |
| 2008 | 0.54 | 0.11-1.12 | 37.77 | 120 | 0.71 | 52 |
| 2009 | 0.88 | 0.52-1.33 | 16.99 | 120 | 1.50 | 52 |

Table 10. Blue catfish age 1+ indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. Mean | 95\% C.I.'s | C.V. | N | River Only (RO) | N |
| 1988 | 0.02 | 0.00-0.05 | 100.00 | 52 | 0.05 | 52 |
| 1989 | 0.10 | 0.00-0.28 | 82.29 | 51 | 0.10 | 52 |
| 1990 | 0.26 | 0.10-0.45 | 29.18 | 52 | 0.61 | 52 |
| 1991 | 0.85 | 0.51-1.28 | 16.69 | 52 | 0.42 | 52 |
| 1992 | 1.22 | 0.71-1.87 | 16.17 | 52 | 0.84 | 52 |
| 1993 | 0.47 | 0.06-1.03 | 42.59 | 52 | 0.57 | 52 |
| 1994 | 0.52 | 0.16-0.99 | 32.59 | 52 | 1.03 | 52 |
| 1995 | 0.48 | 0.14-0.93 | 33.56 | 109 | 0.62 | 52 |
| 1996 | 1.38 | 0.62-2.49 | 22.11 | 120 | 2.33 | 52 |
| 1997 | 3.85 | 2.41-5.89 | 11.17 | 120 | 4.41 | 52 |
| 1998 | 1.99 | 0.95-3.59 | 19.57 | 120 | 3.34 | 52 |
| 1999 | 1.06 | 0.54-1.75 | 19.96 | 114 | 1.73 | 52 |
| 2000 | 0.88 | 0.33-1.65 | 27.38 | 120 | 0.89 | 52 |
| 2001 | 0.55 | 0.30-0.85 | 20.39 | 120 | 0.98 | 52 |
| 2002 | 0.96 | 0.42-1.70 | 23.81 | 120 | 0.84 | 52 |
| 2003 | 1.81 | 0.94-3.08 | 18.02 | 120 | 2.38 | 52 |
| 2004 | 2.62 | 1.78-3.70 | 10.23 | 120 | 4.99 | 52 |
| 2005 | 5.96 | 3.48-9.81 | 11.36 | 120 | 13.52 | 52 |
| 2006 | 4.43 | 2.57-7.24 | 12.36 | 120 | 14.70 | 52 |
| 2007 | 4.79 | 2.80-7.81 | 11.97 | 120 | 7.00 | 52 |
| 2008 | 2.89 | 1.84-4.32 | 11.52 | 120 | 5.85 | 52 |
| 2009 | 3.03 | 1.61-5.23 | 15.61 | 120 | 4.46 | 52 |

Table 11. Channel catfish juvenile indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | River Only |  |
|  | Mean | 95\% C.I.'s |  |  | (RO) | N |
| 1988 | 0.02 | 0.00-0.06 | 85.43 | 52 | 0.03 | 52 |
| 1989 | 1.92 | 1.03-3.22 | 17.10 | 51 | 1.27 | 52 |
| 1990 | 0.04 | 0.00-0.01 | 72.68 | 52 | 0.09 | 52 |
| 1991 | 0.03 | 0.00-0.09 | 100.00 | 52 | 0.02 | 52 |
| 1992 | 0.00 | 0 |  | 52 | 0.00 | 52 |
| 1993 | 0.04 | 0.00-0.12 | 77.30 | 52 | 0.08 | 52 |
| 1994 | 0.05 | 0.00-0.11 | 58.60 | 52 | 0.09 | 52 |
| 1995 | 0.22 | 0.07-0.40 | 33.76 | 109 | 0.40 | 52 |
| 1996 | 0.13 | 0.02-0.26 | 43.48 | 120 | 0.24 | 52 |
| 1997 | 0.05 | 0.00-0.12 | 63.47 | 120 | 0.03 | 52 |
| 1998 | 0.06 | 0.00-0.12 | 49.85 | 120 | 0.04 | 52 |
| 1999 | 0.00 | 0 |  | 114 | 0.00 | 52 |
| 2000 | 0.01 | 0.00-0.02 | 42.25 | 120 | 0.04 | 52 |
| 2001 | 0.00 | 0.00-0.01 | 100.00 | 120 | 0.00 | 52 |
| 2002 | 0.00 | 0.00-0.01 | 100.00 | 120 | 0.00 | 52 |
| 2003 | 0.32 | 0.16-0.50 | 23.67 | 120 | 0.83 | 52 |
| 2004 | 0.19 | 0.08-0.32 | 28.85 | 120 | 0.39 | 52 |
| 2005 | 0.02 | 0.00-0.05 | 56.41 | 120 | 0.09 | 52 |
| 2006 | 0.01 | 0.00-0.02 | 83.74 | 120 | 0.00 | 52 |
| 2007 | 0.00 | 0.00-0.01 | 100.00 | 120 | 0.00 | 52 |
| 2008 | 0.00 | 0 |  | 120 | 0.00 | 52 |
| 2009 | 0.01 | 0.00-0.02 | 100.00 | 120 | 0.02 | 52 |

Table 12. Channel catfish age $1+$ indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | River Onl |  |
|  | Mean | 95\% C.I.'s | C.V. | N | (RO) | N |
| 1988 | 1.41 | 1.05-1.82 | 9.09 | 52 | 0.91 | 52 |
| 1989 | 1.10 | 0.52-1.91 | 21.82 | 51 | 1.20 | 52 |
| 1990 | 2.67 | 1.79-3.83 | 10.56 | 52 | 1.52 | 52 |
| 1991 | 3.70 | 2.47-5.36 | 9.78 | 52 | 1.73 | 52 |
| 1992 | 2.12 | 1.46-2.96 | 10.47 | 52 | 1.48 | 52 |
| 1993 | 0.83 | 0.20-1.80 | 35.01 | 52 | 1.15 | 52 |
| 1994 | 0.84 | 0.50-1.27 | 17.04 | 52 | 1.49 | 52 |
| 1995 | 0.69 | 0.39-1.05 | 18.45 | 109 | 0.58 | 52 |
| 1996 | 1.08 | 0.60-1.71 | 17.84 | 120 | 1.17 | 52 |
| 1997 | 0.84 | 0.47-1.30 | 18.21 | 120 | 1.06 | 52 |
| 1998 | 0.79 | 0.46-1.19 | 17.60 | 120 | 0.68 | 52 |
| 1999 | 0.33 | 0.13-0.56 | 28.23 | 114 | 0.77 | 52 |
| 2000 | 0.25 | 0.11-0.41 | 26.84 | 120 | 0.31 | 52 |
| 2001 | 0.17 | 0.04-0.33 | 38.79 | 120 | 0.16 | 52 |
| 2002 | 0.37 | 0.16-0.61 | 26.31 | 120 | 0.36 | 52 |
| 2003 | 0.28 | 0.15-0.44 | 22.78 | 120 | 0.37 | 52 |
| 2004 | 0.32 | 0.14-0.53 | 26.58 | 120 | 0.54 | 52 |
| 2005 | 0.28 | 0.10-0.49 | 30.49 | 120 | 0.31 | 52 |
| 2006 | 0.21 | 0.10-0.33 | 25.62 | 120 | 0.28 | 52 |
| 2007 | 0.06 | 0.02-0.11 | 32.35 | 120 | 0.14 | 52 |
| 2008 | 0.06 | 0.02-0.10 | 32.99 | 120 | 0.10 | 52 |
| 2009 | 0.06 | 0.00-0.13 | 57.37 | 120 | 0.08 | 52 |

Table 13. Scup indices (1988-2008).

|  | Random Stratified Index (RSI) |  |  |  |
| :---: | :---: | :---: | :---: | ---: |
| Year | Geo. | Mean | $95 \%$ C.I.'s | C.V. |
| 1988 | N |  |  |  |
| 1989 | 4.06 | $2.05-4.41$ | 10.20 | 112 |
| 1990 | 1.90 | $3.14-7.45$ | 10.03 | 112 |
| 1991 | 0.65 | $0.41-2.99$ | 14.99 | 103 |
| 1992 | 3.36 | $2.16-5.01$ | 15.67 | 104 |
| 1993 | 0.90 | $0.53-1.35$ | 10.90 | 104 |
| 1994 | 0.39 | $0.21-0.59$ | 21.36 | 104 |
| 1995 | 0.54 | $0.29-0.83$ | 20.37 | 104 |
| 1996 | 0.21 | $0.09-0.35$ | 28.00 | 104 |
| 1997 | 0.50 | $0.27-0.75$ | 19.83 | 79 |
| 1998 | 0.27 | $0.06-0.52$ | 37.91 | 88 |
| 1999 | 0.13 | $0.02-0.25$ | 41.14 | 105 |
| 2000 | 1.34 | $0.88-1.90$ | 12.80 | 111 |
| 2001 | 0.24 | $0.11-0.37$ | 24.52 | 64 |
| 2002 | 0.96 | $0.58-1.42$ | 15.89 | 104 |
| 2003 | 0.46 | $0.28-0.67$ | 17.38 | 104 |
| 2004 | 1.11 | $0.71-1.59$ | 13.89 | 104 |
| 2005 | 1.58 | $0.99-2.36$ | 13.77 | 104 |
| 2006 | 2.99 | $2.07-4.19$ | 9.47 | 104 |
| 2007 | 0.20 | $0.09-0.31$ | 25.12 | 104 |
| 2008 | 2.97 | $2.07-4.13$ | 9.28 | 104 |

Table 14. Silver perch indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | Bay \& River (BRI) | River Only |  |  |
|  | Mean | 95\% C.I.'s |  |  |  | N | (RO) | N |
| 1988 | 0.61 | 0.35-0.92 | 18.30 | 172 | 0.65 | 172 | 1.02 | 65 |
| 1989 | 0.53 | 0.33-0.76 | 16.32 | 189 | 0.56 | 189 | 1.63 | 63 |
| 1990 | 0.69 | 0.49-0.92 | 11.94 | 185 | 0.75 | 185 | 4.08 | 59 |
| 1991 | 0.35 | 0.21-0.51 | 17.33 | 179 | 0.40 | 179 | 1.47 | 62 |
| 1992 | 0.81 | 0.49-1.18 | 15.80 | 178 | 0.86 | 178 | 1.95 | 61 |
| 1993 | 0.45 | 0.29-0.63 | 16.01 | 180 | 0.45 | 180 | 0.60 | 63 |
| 1994 | 0.25 | 0.11-0.40 | 25.42 | 180 | 0.26 | 180 | 0.37 | 63 |
| 1995 | 0.58 | 0.34-0.87 | 15.65 | 180 | 0.65 | 180 | 1.81 | 67 |
| 1996 | 0.59 | 0.38-0.84 | 15.63 | 304 | 0.58 | 183 | 1.18 | 66 |
| 1997 | 0.71 | 0.50-0.94 | 12.07 | 316 | 0.79 | 192 | 1.43 | 75 |
| 1998 | 0.24 | 0.15-0.33 | 16.77 | 316 | 0.24 | 192 | 0.53 | 75 |
| 1999 | 0.70 | 0.49-0.94 | 12.42 | 309 | 0.74 | 186 | 2.51 | 75 |
| 2000 | 0.68 | 0.46-0.93 | 13.56 | 317 | 0.76 | 192 | 2.12 | 74 |
| 2001 | 0.70 | 0.47-0.97 | 13.77 | 327 | 0.85 | 200 | 3.17 | 75 |
| 2002 | 0.44 | 0.24-0.67 | 20.16 | 269 | 0.41 | 146 | 1.67 | 75 |
| 2003 | 0.63 | 0.40-0.90 | 15.49 | 315 | 0.66 | 192 | 0.71 | 75 |
| 2004 | 0.34 | 0.22-0.48 | 16.50 | 315 | 0.36 | 192 | 0.80 | 75 |
| 2005 | 0.76 | 0.52-1.03 | 12.64 | 315 | 0.77 | 192 | 2.20 | 75 |
| 2006 | 1.21 | 0.84-1.64 | 11.31 | 283 | 1.22 | 174 | 4.45 | 67 |
| 2007 | 0.75 | 0.50-1.03 | 13.53 | 315 | 0.68 | 192 | 2.26 | 75 |
| 2008 | 0.49 | 0.34-0.66 | 13.31 | 315 | 0.46 | 192 | 0.84 | 75 |
| 2009 | 1.00 | 0.72-1.32 | 10.83 | 315 | 0.92 | 192 | 1.74 | 75 |

Table 15. Spot indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | Bay \& River (BRI) | River Only |  |  |
|  | Mean | 95\% C.I.'s |  |  |  | N | (RO) | N |
| 1988 | 67.01 | 46.36-96.67 | 4.29 | 231 | 67.45 | 231 | 50.20 | 84 |
| 1989 | 31.41 | 24.51-40.18 | 3.44 | 252 | 32.27 | 252 | 54.19 | 84 |
| 1990 | 44.78 | 32.34-61.85 | 4.14 | 248 | 45.28 | 248 | 53.06 | 81 |
| 1991 | 16.83 | 12.28-21.60 | 4.66 | 238 | 16.56 | 238 | 21.44 | 83 |
| 1992 | 1.92 | 1.45-2.49 | 8.20 | 238 | 1.96 | 238 | 4.39 | 82 |
| 1993 | 9.78 | 7.23-13.13 | 5.68 | 240 | 9.74 | 240 | 11.85 | 84 |
| 1994 | 9.23 | 6.88-12.27 | 5.61 | 240 | 9.07 | 240 | 8.88 | 84 |
| 1995 | 1.56 | 1.15-2.05 | 9.25 | 248 | 1.52 | 248 | 2.37 | 92 |
| 1996 | 5.26 | 4.15-6.60 | 5.30 | 407 | 4.52 | 244 | 4.84 | 88 |
| 1997 | 11.50 | 9.11-14.45 | 4.20 | 421 | 8.63 | 256 | 19.68 | 100 |
| 1998 | 2.51 | 1.92-3.23 | 7.36 | 374 | 1.88 | 214 | 3.04 | 96 |
| 1999 | 4.72 | 3.63-6.07 | 6.07 | 402 | 3.98 | 238 | 6.61 | 100 |
| 2000 | 3.32 | 2.57-4.23 | 6.51 | 421 | 2.70 | 253 | 4.94 | 97 |
| 2001 | 3.09 | 2.45-3.85 | 6.06 | 432 | 2.83 | 264 | 3.69 | 100 |
| 2002 | 2.89 | 2.10-3.88 | 8.38 | 360 | 2.09 | 196 | 3.12 | 100 |
| 2003 | 2.85 | 2.25-3.56 | 6.32 | 420 | 2.58 | 256 | 2.32 | 100 |
| 2004 | 3.96 | 3.14-4.95 | 5.68 | 420 | 3.21 | 255 | 6.91 | 99 |
| 2005 | 12.12 | 9.80-14.94 | 3.78 | 420 | 8.91 | 256 | 16.58 | 100 |
| 2006 | 3.37 | 2.71-4.16 | 5.61 | 420 | 2.67 | 256 | 3.20 | 100 |
| 2007 | 9.17 | 7.38-11.35 | 4.18 | 420 | 7.79 | 256 | 12.75 | 100 |
| 2008 | 19.89 | 15.16-26.01 | 4.22 | 420 | 16.83 | 256 | 16.77 | 100 |
| 2009 | 6.08 | 4.96-7.40 | 4.39 | 420 | 4.74 | 256 | 9.05 | 100 |

Table 16. Striped bass indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N |  |  |
|  | Mean | 95\% C.I.'s |  |  | $\begin{aligned} & \hline \text { River Only } \\ & \text { (RO) } \end{aligned}$ | N |
| 1988 | 1.24 | 0.65-2.06 | 19.19 | 35 | 1.93 | 35 |
| 1989 | 1.65 | 1.12-2.32 | 11.51 | 37 | 1.59 | 37 |
| 1990 | 1.06 | 0.49-1.84 | 22.33 | 36 | 1.14 | 36 |
| 1991 | 1.09 | 0.31-2.33 | 31.00 | 36 | 1.02 | 36 |
| 1992 | 1.22 | 0.76-1.81 | 13.18 | 39 | 2.15 | 39 |
| 1993 | 2.52 | 1.09-4.94 | 19.32 | 41 | 3.30 | 41 |
| 1994 | 1.31 | 0.85-1.87 | 12.58 | 39 | 1.07 | 39 |
| 1995 | 0.63 | 0.34-0.99 | 20.19 | 61 | 1.22 | 39 |
| 1996 | 0.61 | 0.32-0.95 | 20.56 | 90 | 1.19 | 39 |
| 1997 | 0.55 | 0.25-0.93 | 24.75 | 90 | 0.41 | 39 |
| 1998 | 0.89 | 0.44-1.47 | 21.30 | 90 | 1.22 | 39 |
| 1999 | 0.21 | 0.00-0.47 | 51.55 | 84 | 0.26 | 39 |
| 2000 | 1.54 | 0.76-2.67 | 19.70 | 90 | 2.72 | 39 |
| 2001 | 0.53 | 0.27-0.85 | 21.84 | 90 | 1.94 | 39 |
| 2002 | 0.71 | 0.42-1.07 | 17.34 | 90 | 1.68 | 39 |
| 2003 | 0.63 | 0.24-1.13 | 27.59 | 90 | 1.01 | 39 |
| 2004 | 0.33 | 0.17-0.52 | 22.68 | 90 | 0.45 | 39 |
| 2005 | 0.59 | 0.30-0.95 | 21.79 | 90 | 0.53 | 39 |
| 2006 | 0.27 | 0.13-0.42 | 23.65 | 90 | 0.55 | 39 |
| 2007 | 0.37 | 0.21-0.55 | 20.10 | 90 | 0.74 | 39 |
| 2008 | 0.62 | 0.22-1.15 | 29.31 | 90 | 1.58 | 39 |
| 2009 | 0.48 | 0.28-0.70 | 17.93 | 90 | 1.06 | 39 |

Table 17. Summer flounder indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | Bay \& River (BRI) | River Only |  |  |
|  | Mean | 95\% C.I.'s |  |  |  | N | (RO) | N |
| 1988 | 0.54 | 0.35-0.75 | 14.99 | 143 | 0.53 | 143 | 0.54 | 36 |
| 1989 | 1.24 | 0.94-1.58 | 8.77 | 162 | 1.23 | 162 | 0.96 | 36 |
| 1990 | 2.54 | 2.06-3.09 | 5.73 | 162 | 2.54 | 162 | 2.61 | 36 |
| 1991 | 2.79 | 2.26-3.41 | 5.66 | 153 | 2.78 | 153 | 1.42 | 36 |
| 1992 | 0.92 | 0.70-1.17 | 9.25 | 153 | 0.91 | 153 | 0.49 | 36 |
| 1993 | 0.52 | 0.38-0.68 | 11.87 | 153 | 0.53 | 153 | 0.49 | 36 |
| 1994 | 2.54 | 2.01-3.15 | 6.39 | 153 | 2.50 | 153 | 1.08 | 36 |
| 1995 | 0.71 | 0.52-0.92 | 10.89 | 149 | 0.72 | 149 | 0.74 | 36 |
| 1996 | 0.81 | 0.62-1.02 | 9.32 | 224 | 0.86 | 153 | 0.62 | 36 |
| 1997 | 0.89 | 0.69-1.12 | 8.77 | 226 | 0.97 | 153 | 0.70 | 36 |
| 1998 | 0.73 | 0.55-0.93 | 9.92 | 226 | 0.78 | 153 | 0.17 | 36 |
| 1999 | 0.53 | 0.41-0.67 | 9.94 | 219 | 0.58 | 147 | 0.36 | 36 |
| 2000 | 0.57 | 0.43-0.73 | 10.81 | 227 | 0.62 | 154 | 0.52 | 36 |
| 2001 | 0.47 | 0.34-0.61 | 11.84 | 236 | 0.52 | 161 | 0.53 | 36 |
| 2002 | 0.77 | 0.54-1.04 | 12.21 | 179 | 0.80 | 107 | 0.43 | 36 |
| 2003 | 0.44 | 0.33-0.56 | 10.95 | 225 | 0.43 | 153 | 0.50 | 36 |
| 2004 | 1.30 | 1.03-1.60 | 7.50 | 225 | 1.40 | 153 | 1.17 | 36 |
| 2005 | 0.35 | 0.25-0.46 | 13.18 | 225 | 0.36 | 153 | 0.29 | 36 |
| 2006 | 0.80 | 0.60-1.02 | 10.03 | 203 | 0.87 | 139 | 0.59 | 32 |
| 2007 | 1.00 | 0.78-1.24 | 8.22 | 225 | 1.04 | 153 | 0.53 | 36 |
| 2008 | 1.35 | 1.10-1.63 | 6.68 | 225 | 1.49 | 153 | 1.09 | 36 |
| 2009 | 0.75 | 0.58-0.92 | 8.76 | 225 | 0.82 | 153 | 0.84 | 36 |

Table 18. Weakfish indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  | C.V. | N | Bay \& Rive (BRI) | River Only |  |  |
|  | Mean | 95\% C.I.'s |  |  |  | N | (RO) | N |
| 1988 | 8.13 | 5.37-12.07 | 8.12 | 173 | 8.89 | 173 | 21.72 | 63 |
| 1989 | 11.74 | 8.18-16.88 | 6.44 | 189 | 12.22 | 189 | 21.27 | 63 |
| 1990 | 4.46 | 3.10-6.26 | 8.44 | 184 | 4.87 | 184 | 30.01 | 59 |
| 1991 | 3.16 | 2.32-4.21 | 7.92 | 179 | 3.56 | 179 | 15.32 | 62 |
| 1992 | 6.78 | 4.74-9.53 | 7.39 | 178 | 6.93 | 178 | 15.91 | 61 |
| 1993 | 5.81 | 4.06-8.17 | 7.76 | 180 | 6.12 | 180 | 15.42 | 63 |
| 1994 | 2.51 | 1.76-3.47 | 9.59 | 180 | 2.67 | 180 | 7.04 | 63 |
| 1995 | 5.95 | 4.26-8.18 | 7.20 | 186 | 6.07 | 186 | 11.00 | 69 |
| 1996 | 7.26 | 5.33-9.78 | 6.31 | 305 | 7.85 | 183 | 7.42 | 66 |
| 1997 | 6.81 | 5.26-8.74 | 5.38 | 316 | 7.15 | 192 | 14.82 | 75 |
| 1998 | 7.60 | 5.46-10.45 | 6.65 | 269 | 8.18 | 150 | 9.95 | 71 |
| 1999 | 6.78 | 5.01-9.06 | 6.28 | 303 | 7.38 | 180 | 16.25 | 75 |
| 2000 | 8.35 | 6.34-10.92 | 5.42 | 316 | 9.39 | 191 | 11.09 | 74 |
| 2001 | 5.09 | 3.74-6.82 | 6.93 | 327 | 5.14 | 200 | 11.52 | 75 |
| 2002 | 6.93 | 4.27-10.94 | 9.89 | 270 | 6.30 | 147 | 8.59 | 75 |
| 2003 | 9.23 | 6.72-12.54 | 6.04 | 315 | 9.34 | 192 | 5.42 | 75 |
| 2004 | 6.66 | 4.94-8.88 | 6.24 | 315 | 7.27 | 192 | 10.47 | 75 |
| 2005 | 5.69 | 4.26-7.50 | 6.31 | 315 | 5.93 | 192 | 7.10 | 75 |
| 2006 | 6.34 | 4.83-8.25 | 5.80 | 315 | 6.21 | 192 | 6.20 | 75 |
| 2007 | 5.35 | 3.99-7.08 | 6.51 | 315 | 5.30 | 192 | 14.37 | 75 |
| 2008 | 5.77 | 4.33-7.60 | 6.26 | 315 | 5.51 | 192 | 25.87 | 75 |
| 2009 | 6.18 | 4.75-7.96 | 5.63 | 315 | 6.25 | 192 | 11.44 | 75 |

Table 19. White catfish juvenile indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | River Onl |  |
|  | Mean | 95\% C.I.'s | C.V. | N | (RO) | N |
| 1988 | 0.25 | 0.11-0.41 | 26.68 | 52 | 0.61 | 52 |
| 1989 | 3.63 | 2.01-6.12 | 14.03 | 51 | 3.33 | 52 |
| 1990 | 0.76 | 0.57-0.97 | 9.89 | 52 | 0.82 | 52 |
| 1991 | 0.07 | 0.02-0.12 | 34.21 | 52 | 0.19 | 52 |
| 1992 | 0.81 | 0.62-1.01 | 9.08 | 52 | 0.50 | 52 |
| 1993 | 0.76 | 0.39-1.24 | 21.11 | 52 | 1.14 | 52 |
| 1994 | 0.26 | 0.19-0.32 | 11.29 | 52 | 0.34 | 52 |
| 1995 | 0.21 | 0.08-0.35 | 29.33 | 109 | 0.46 | 52 |
| 1996 | 0.36 | 0.18-0.55 | 22.23 | 120 | 1.18 | 53 |
| 1997 | 0.37 | 0.23-0.53 | 17.47 | 120 | 0.94 | 52 |
| 1998 | 0.07 | 0.04-0.10 | 22.96 | 120 | 0.34 | 52 |
| 1999 | 0.00 | 0.00-0.01 | 100.00 | 114 | 0.00 | 52 |
| 2000 | 0.05 | 0.00-0.12 | 58.53 | 120 | 0.09 | 52 |
| 2001 | 0.02 | 0.00-0.04 | 73.60 | 120 | 0.03 | 52 |
| 2002 | 0.00 | 0 |  | 120 | 0.00 | 52 |
| 2003 | 0.29 | 0.17-0.42 | 19.28 | 120 | 0.99 | 52 |
| 2004 | 0.12 | 0.04-0.20 | 33.23 | 120 | 0.19 | 52 |
| 2005 | 0.04 | 0.00-0.09 | 44.35 | 120 | 0.18 | 52 |
| 2006 | 0.08 | 0.03-0.14 | 32.32 | 120 | 0.18 | 52 |
| 2007 | 0.02 | 0.00-0.04 | 46.28 | 120 | 0.07 | 52 |
| 2008 | 0.03 | 0.00-0.08 | 80.25 | 120 | 0.02 | 52 |
| 2009 | 0.13 | 0.07-0.20 | 23.87 | 120 | 0.36 | 52 |

Table 20. White catfish age $1+$ indices (1988-2009).

| Year | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | River Only |  |
|  | Mean | 95\% C.I.'s | C.V. | N | (RO) | N |
| 1988 | 1.88 | 1.29-2.62 | 10.81 | 52 | 3.16 | 52 |
| 1989 | 3.23 | 1.68-5.67 | 15.78 | 51 | 4.35 | 52 |
| 1990 | 3.46 | 2.13-5.34 | 11.82 | 52 | 6.75 | 52 |
| 1991 | 2.12 | 0.90-4.11 | 21.68 | 52 | 2.31 | 52 |
| 1992 | 4.27 | 3.41-5.31 | 5.38 | 52 | 3.97 | 52 |
| 1993 | 2.13 | 1.12-3.64 | 17.19 | 52 | 1.66 | 52 |
| 1994 | 1.79 | 1.23-2.48 | 10.86 | 52 | 2.72 | 52 |
| 1995 | 0.94 | 0.45-1.61 | 22.21 | 109 | 1.77 | 52 |
| 1996 | 1.05 | 0.76-1.40 | 10.78 | 120 | 3.11 | 53 |
| 1997 | 1.85 | 1.32-2.49 | 9.82 | 120 | 3.45 | 52 |
| 1998 | 1.21 | 0.76-1.77 | 14.40 | 120 | 2.45 | 52 |
| 1999 | 0.56 | 0.36-0.79 | 15.31 | 114 | 1.51 | 52 |
| 2000 | 0.29 | 0.15-0.45 | 22.91 | 120 | 0.66 | 52 |
| 2001 | 0.29 | 0.14-0.47 | 24.65 | 120 | 0.54 | 52 |
| 2002 | 0.36 | 0.11-0.66 | 33.57 | 120 | 0.52 | 52 |
| 2003 | 0.48 | 0.26-0.74 | 20.34 | 120 | 1.13 | 52 |
| 2004 | 0.28 | 0.15-0.42 | 21.66 | 120 | 0.66 | 52 |
| 2005 | 0.41 | 0.23-0.62 | 19.83 | 120 | 0.98 | 52 |
| 2006 | 0.34 | 0.22-0.11 | 16.10 | 120 | 0.73 | 52 |
| 2007 | 0.23 | 0.11-0.35 | 24.29 | 120 | 0.62 | 52 |
| 2008 | 0.20 | 0.10-0.30 | 23.09 | 120 | 0.55 | 52 |
| 2009 | 0.19 | 0.09-0.29 | 24.47 | 120 | 0.46 | 52 |

Table 21. White perch juvenile indices (1988-2009).

|  | Random Stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Geo. <br> Mean | 95\% C.I.'s | C.V. | N | River Only (RO) | N |
| 1988 | 6.15 | 3.68-9.91 | 10.75 | 35 | 5.29 | 35 |
| 1989 | 12.93 | 6.69-24.25 | 11.29 | 37 | 13.33 | 37 |
| 1990 | 3.24 | 1.84-5.32 | 13.89 | 36 | 3.31 | 36 |
| 1991 | 4.15 | 1.35-10.30 | 23.95 | 36 | 2.30 | 36 |
| 1992 | 1.71 | 0.90-2.86 | 17.70 | 39 | 1.21 | 39 |
| 1993 | 19.24 | 5.55-61.55 | 18.75 | 41 | 17.91 | 41 |
| 1994 | 14.28 | 7.85-25.39 | 10.02 | 39 | 8.43 | 39 |
| 1995 | 2.58 | 1.34-4.48 | 16.65 | 61 | 4.61 | 39 |
| 1996 | 24.41 | 12.94-45.29 | 9.27 | 90 | 22.69 | 39 |
| 1997 | 9.34 | 6.04-14.19 | 8.22 | 90 | 10.00 | 39 |
| 1998 | 3.84 | 1.98-6.86 | 15.38 | 90 | 7.13 | 39 |
| 1999 | 0.74 | 0.39-1.19 | 20.54 | 84 | 2.38 | 39 |
| 2000 | 8.23 | 4.01-15.99 | 13.74 | 90 | 16.90 | 39 |
| 2001 | 1.93 | 0.95-3.39 | 18.83 | 90 | 5.99 | 39 |
| 2002 | 4.66 | 3.47-6.16 | 6.77 | 90 | 9.48 | 39 |
| 2003 | 21.98 | 9.91-47.40 | 11.89 | 90 | 15.70 | 39 |
| 2004 | 6.52 | 3.27-12.26 | 14.05 | 90 | 4.32 | 39 |
| 2005 | 11.75 | 6.70-20.11 | 9.91 | 90 | 8.39 | 39 |
| 2006 | 4.35 | 2.52-7.14 | 12.48 | 90 | 5.00 | 39 |
| 2007 | 3.10 | 1.43-5.93 | 18.56 | 90 | 4.62 | 39 |
| 2008 | 8.39 | 3.40-19.04 | 16.94 | 90 | 9.74 | 39 |
| 2009 | 7.22 | 3.95-12.64 | 12.02 | 90 | 9.03 | 39 |

Table 22. White perch age $1+$ indices (1988-2009).

| Year | Random stratified Index (RSI) |  |  |  | Original Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Geo. |  |  |  | River O |  |
|  | Mean | 95\% C.I.'s | C.V. | N | (RO) | N |
| 1988 | 39.57 | 26.69-58.42 | 5.15 | 46 | 35.10 | 46 |
| 1989 | 22.78 | 16.00-32.25 | 5.29 | 46 | 25.86 | 46 |
| 1990 | 35.39 | 21.9-56.83 | 6.44 | 45 | 31.97 | 45 |
| 1991 | 39.52 | 28.45-54.75 | 4.31 | 44 | 29.49 | 44 |
| 1992 | 11.94 | 7.73-18.17 | 7.68 | 48 | 15.77 | 48 |
| 1993 | 10.74 | 4.89-22.37 | 13.98 | 50 | 15.04 | 50 |
| 1994 | 24.61 | 15.49-38.78 | 6.79 | 48 | 18.77 | 48 |
| 1995 | 13.69 | 6.17-29.12 | 13.35 | 70 | 40.82 | 48 |
| 1996 | 9.03 | 5.29-15.00 | 10.13 | 116 | 13.27 | 49 |
| 1997 | 19.37 | 10.56-34.90 | 9.40 | 120 | 20.25 | 52 |
| 1998 | 10.89 | 6.70-17.36 | 8.78 | 120 | 27.44 | 52 |
| 1999 | 10.34 | 5.97-17.46 | 10.03 | 114 | 23.48 | 52 |
| 2000 | 7.65 | 3.79-14.63 | 13.72 | 120 | 17.31 | 52 |
| 2001 | 4.62 | 2.54-7.92 | 13.36 | 120 | 17.09 | 52 |
| 2002 | 7.22 | 4.99-10.28 | 7.51 | 120 | 20.61 | 52 |
| 2003 | 19.13 | 9.95-36.00 | 10.14 | 120 | 27.35 | 52 |
| 2004 | 6.84 | 3.83-11.72 | 11.76 | 120 | 8.71 | 52 |
| 2005 | 8.40 | 5.30-13.04 | 8.95 | 120 | 9.34 | 52 |
| 2006 | 6.45 | 4.09-9.89 | 9.48 | 120 | 13.11 | 52 |
| 2007 | 6.48 | 3.88-10.46 | 10.61 | 120 | 19.58 | 52 |
| 2008 | 5.08 | 2.51-9.53 | 15.20 | 120 | 17.27 | 52 |
| 2009 | 7.61 | 4.63-12.17 | 9.87 | 120 | 20.68 | 52 |

## FIGURES

Figure 1. The VIMS trawl survey random stratified design in the Chesapeake Bay. Transect lines indicate geographic regions as designated below.

| Chesapeake Bay | B1 | Bottom Bay |
| :---: | :---: | :---: |
|  | B2 | Lower Bay |
|  | B3 | Upper Bay |
| James River | J1 | Bottom James |
|  | J2 | Lower James |
|  | J3 | Upper James |
|  | J4 | Top James |
| York River | Y1 | Bottom York |
|  | Y2 | Lower York |
|  | Y3 | Upper York |
|  | Y4 | Top York (lower Pamunkey River) |
| Rappahannock River | R1 | Bottom Rappahannock |
|  | R2 | Lower Rappahannock |
|  | R3 | Upper Rappahannock |
|  | R4 | Top Rappahannock |
| Mobjack Bay | MB | Routine monitoring established March 2010 |



Figure 1 (continued)


Figure 2. American eel random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the time series average based on the RSI (Top), and distribution of American eel (all age classes combined) from index strata and months (Bottom).


Figure 3. Spring juvenile Atlantic croaker random stratified (RSI, 95\% C.I.), fixed transect (Rivers only - RO), and Bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of Atlantic croaker juveniles from index strata and months (Bottom).


Figure 4. Bay anchovy random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the time series average based on the RSI (Top), and distribution of bay anchovy from index strata and months (Bottom).


Figure 5. Black sea bass random stratified index (RSI, 95\% C.I.) and fixed transect (Rivers only RO ), and bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of juvenile black sea bass from index strata and months (Bottom).


Figure 6. Juvenile blue catfish random stratified index (RSI, 95\% C.I.) and fixed transect (Rivers only -RO) indices and the time series average based on the RSI (Top), and distribution of juvenile blue catfish from index strata and months (Bottom).


Figure 7. Age $1+$ blue catfish random stratified index (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the time series average based on the RSI (Top), and distribution of age $1+$ blue catfish from index strata and months (Bottom).


Figure 8. Juvenile channel catfish random stratified index (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the time series average based on the RSI (Top), and distribution of juvenile channel catfish from index strata and months (Bottom). One juvenile channel catfish was captured in this time period.


Figure 9. Age $1+$ channel catfish random stratified index (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the time series average based on the RSI (Top), and distribution of age $1+$ channel catfish from index strata and months (Bottom).


Figure 10. Juvenile scup random stratified index (RSI, 95\% C.I.) and the time series average (Top), and distribution of scup from index strata and months (Bottom).


Figure 11. Juvenile silver perch random stratified (RSI, 95\% C.I.), fixed transect (Rivers only RO), and Bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of juvenile silver perch from index strata and months (Bottom).


Figure 12. Juvenile spot random stratified (RSI, 95\% C.I.) fixed transect (Rivers only - RO), and Bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of juvenile spot from index strata and months (Bottom).


Figure 13. Juvenile striped bass random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only -RO ) indices and the time series average based on the RSI (Top), and distribution of juvenile striped bass from index strata and months (Bottom).


Figure 14. Juvenile summer flounder random stratified (RSI, 95\% C.I.), fixed transect (Rivers only - RO), and Bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of juvenile summer flounder from index strata and months (Bottom).


Figure 15. Juvenile weakfish random stratified (RSI, 95\% C.I.), fixed transect (Rivers only RO), and Bay and fixed river station (BRI) indices and the time series average based on the RSI (Top), and distribution of juvenile weakfish from index strata and months (Bottom).


Figure 16. Juvenile white catfish random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only - RO) indices and the times series average based on the RSI (Top), and distribution of juvenile white catfish from index strata and months (Bottom).


Figure 17. White catfish age $1+$ random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only -RO ) indices and the time series average based on the RSI (Top), and distribution of age $1+$ white catfish from index strata and months (Bottom).


Figure 18. Juvenile white perch random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only -RO ) indices and the time series average based on the RSI (Top), and distribution of juvenile white perch from index strata and months (Bottom).


Figure 19. White perch age $1+$ random stratified (RSI, 95\% C.I.) and fixed transect (Rivers only -RO ) indices and the time series average based on the RSI (Top), and distribution of age $1+$ white perch from index strata and months (Bottom).

Appendix Table 1. Trawl survey advisory requests, data requests, and specimen requests from June 2009 to May 2010.

| Contact | Agency | Nature of Request |
| :--- | :--- | :--- |
| Pat McGrath | VIMS | Gar specimens |
| Karin Limburg | SUNY | Alosine specimens |
| Edward McGinley | WVU | Bay anchow specimens |
| Edward McGinley | WVU | Atlantic croaker specimens |
| Edward McGinley | WVU | Atlantic menhaden specimens |
| Joel Boehm | River Project | Dusky pipefish specimens |
| Joel Boehm | River Project | Northern pipefish specimens |
| Joel Boehm | River Project | Lined seahorse specimens |
| Paul Gerdes | VIMS | Naked goby specimens |
| Paul Gerdes | VIMS | Bay anchow specimens |
| Paul Gerdes | VIMS | Striped anchow specimens |
| Paul Gerdes | VIMS | Scup specimens |
| Alicia Nelson | VMRC | 2009 Horseshoe Crab data |
| Sally Upton | VIMS | Fish specimens for education dept |
| Laura Lee | VMRC | Eel data |
| Jack Travelstead | VMRC | Blue catfish data summary |
| Mike Newman | VIMS | Fish for mercury study |
| Laura Lee | VMRC | Spot indices |
| Rob O'Reilly | VMRC | Atlantic Croaker Indices, fall and spring 2008 |
| Eric Hilton | VIMS | Feather blenny specimens |
| Eric Hilton | VIMS | Striped blenny specimens |
| Eric Hilton | VIMS | Crested blenny specimens |
| Troy Tuckey | VIMS | White perch juveniles |
| Lyle Varnell | VIMS | Fish distribution data |
| Lyle Varnell | VIMS | James River water temperature data |
| VIMS Aquarium | VIMS | Live specimens for display and education |
| Tracey Sutton | VIMS | Spiny dogfish data |
| Chris Hager | VA Sea Grant | Atlantic sturgeon data |
| Tom Miller | CBL | Blue crab indices 1968-2009, age 0 spring/fall, age 1+, adult female |
|  |  |  |

Appendix Table 2. Mobjack Bay pooled catch for March to May 2010 from 51 tows.
Adjusted Percent of Catch Excludes Bay Anchovy

| Species | Number of Fish (All) | Frequency | Percent of Catch | Catch <br> Per Trawl | Adjusted Percent of Catch | Number of Fish YOY | Average <br> Length <br> (mm) | Standard Error (length) | Minimum Length (mm) | Maximum Length (mm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| bay anchow | 31,806 | 44 | 91.94 | 623.65 |  | 31,606 | 59 | 0.19 | 33 | 88 |
| spot | 2,065 | 28 | 5.97 | 40.49 | 74.09 | 1,641 | 81 | 2.12 | 19 | 194 |
| weakfish | 246 | 15 | 0.71 | 4.82 | 8.83 | 245 | 166 | 1.25 | 117 | 245 |
| Atlantic croaker | 219 | 17 | 0.63 | 4.29 | 7.86 | 5 | 202 | 1.85 | 97 | 343 |
| Atlantic menhaden | 86 | 15 | 0.25 | 1.69 | 3.09 | 16 | 114 | 4.76 | 29 | 257 |
| spotted hake | 62 | 18 | 0.18 | 1.22 | 2.22 | 62 | 142 | 2.87 | 97 | 180 |
| Atlantic herring | 27 | 7 | 0.08 | 0.53 | 0.97 | . | 51 | 2.35 | 31 | 59 |
| northern pipefish | 12 | 8 | 0.03 | 0.24 | 0.43 | . | 143 | 7.2 | 103 | 169 |
| alewife | 12 | 1 | 0.03 | 0.24 | 0.43 | 12 | 120 | 2.59 | 105 | 134 |
| hogchoker | 11 | 6 | 0.03 | 0.22 | 0.39 | 0 | 105 | 1.96 | 92 | 114 |
| summer flounder | 10 | 8 | 0.03 | 0.2 | 0.36 | 1 | 398 | 42.12 | 89 | 497 |
| silver perch | 6 | 4 | 0.02 | 0.12 | 0.22 | 5 | 137 | 6.35 | 124 | 166 |
| oyster toadfish | 6 | 1 | 0.02 | 0.12 | 0.22 | . | 67 | 9.28 | 49 | 97 |
| Atlantic cutlassfish | 5 | 3 | 0.01 | 0.1 | 0.18 |  | 365 | 22.3 | 329 | 452 |
| butterfish | 4 | 3 | 0.01 | 0.08 | 0.14 | 4 | 63 | 12.61 | 43 | 99 |
| naked goby | 3 | 3 | 0.01 | 0.06 | 0.11 | . | 54 | 3.71 | 47 | 59 |
| kingfish spp | 2 | 2 | 0.01 | 0.04 | 0.07 | 1 | 202 | 62.5 | 139 | 264 |
| lined seahorse | 2 | 2 | 0.01 | 0.04 | 0.07 |  | 77 | 8 | 69 | 85 |
| American shad | 2 | 1 | 0.01 | 0.04 | 0.07 | 2 | 126 | 4 | 122 | 130 |
| blueback herring | 1 | 1 | 0.00 | 0.02 | 0.04 | 0 | 233 | . | 233 | 233 |
| striped bass | 1 | 1 | 0.00 | 0.02 | 0.04 | 0 | 486 | . | 486 | 486 |
| northern puffer | 1 | 1 | 0.00 | 0.02 | 0.04 | 0 | 199 | . | 199 | 199 |
| northern searobin | 1 | 1 | 0.00 | 0.02 | 0.04 | 1 | 51 | . | 51 | 51 |
| striped blenny | 1 | 1 | 0.00 | 0.02 | 0.04 | . | 65 | . | 65 | 65 |
| feather blenny | 1 | 1 | 0.00 | 0.02 | 0.04 | . | 55 | . | 55 | 55 |
| rough scad | 1 | 1 | 0.00 | 0.02 | 0.03 | . | 50 | . | 50 | 50 |

Appendix Figure 1. Length frequency distributions by species from June 2009 to May 2010. Red bars are index-sized fish. (Note that actual indices are calculated using a subset of months and strata. Therefore, not all index-sized fish are included.)



Appendix Figure 1. (continued)



Appendix Figure 1. (continued)



Appendix Figure 1. (continued)



Appendix Figure 1. (continued)



Appendix Figure 1. (continued)



Appendix Figure 1. (continued)


