Soil Physics

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WATER PERMEABILITY OF SOILS AMENDED WITH SEWAGE SLUDGE ON SHORT-ROTATION PLANTATIONS IN EUROPE

Abstract. Soil permeability is the measure of the soil's ability to permit water to flow through soil pores. It depends chiefly on the volume and geometry of soil pores, bulk density, and soil structure. Studied soils were overgrown with short-rotation plantations of *Salix* and amended with sewage sludge in Germany, Estonia, and Poland. Sewage sludge application as well as the use of soils influenced water permeability.

Soil permeability or hydraulic conductivity is the measure of the soil's ability to permit water to flow through soil pores. It depends on the volume and geometry of soil pores, bulk density, and soil structure [3]. Soil water permeability is determined on the basis of K coefficient, which was characterised in Darcy's equation for the discharge velocity of soil water flowing through saturated soil: $v = K * I$, where v is the discharge velocity, i.e. the quantity of water flowing in a unit of time through a unit of gross cross-sectional area of soil at right angles to the direction of flow, and K is the hydraulic conductivity (coefficient of permeability), and I is the decrease in hydraulic pressure calculated

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on the basis of pressure and length of flow. The coefficient of permeability may be mathematically predicted, measured in field conditions or at a laboratory in saturated and unsaturated soils [1, 3, 9].

Soil water permeability affects the availability of water for plants and determines the flow of nutrients in soils [10]. According to Walczak et al. [11] recognition of hydrophysical properties is necessary to interpret and predict most processes occurring in the soil. Measurements of soil saturated hydraulic conductivity and their spatial variations are important to characterise hydrological processes in the field scale [2].

The aim of the study has been to determine whether application of sewage sludge influences the water permeability of soil.

MATERIALS AND METHODS

Studied soils were overgrown with short-rotation plantations of *Salix* and amended with sewage sludge in Germany, Estonia, and Poland. In Germany the experimental field was located in Schönberg-Holstein in the north of Germany, 3 km away from the Baltic Sea and 20 km away from the City of Kiel (geographical coordinates: 54°23'50.55''0N, 10°23'05.89''5E). The annual average precipitation rate in this region was 825 mm with the annual average temperature of 10.8 °C. In Estonia the experimental field was situated in the Tartu Region, Vara Borough, with the coordinates of 58°30'18''8N, 26°55'29''9E. The annual average precipitation in this region was 620 mm, and the average temperature was 5.2 °C. In Poland the field was located in Lisiecice, in the Opole Region, Pawłowice Borough, with the coordinates of 50°15'38''8N, 17°52'36''6E. The annual average precipitation in the region was 650 mm, and the annual temperature was 8.2°C.

The experimental field was a randomised block design, and each sludge treatment was repeated five times. Each block consisted of three sludge treatment processes for one willow clone.

According to the World Reference Base for Soil Resources [12], Polish and German studied soils were classified as *Cambisols*. They contain less than 2.5% of organic carbon (OC) in topsoil and do not meet the requirements of *mollic* horizon. Their sub-surface horizon is *cambic* and their texture is loamy. Estonian soils were classified as *Phaeozems*. They have topsoil rich in organic matter which contains more than 2.5% of OC. Organic matter content, colour and pH enable to classify it as *mollic* diagnostic horizon. Their texture is sandy loam.

The sewage sludge applied on the studied soils contained considerable amounts of organic matter and nutrients (Table 1). Sewage sludge was stabilised with lime, which is a normal practice at sewage treatment plants [4]. Sewage sludge was applied once in two rates calculated on the basis of phosphorus content. Sewage sludge was incorporated and thoroughly mixed by disc harrows

two weeks before willow planting. The single rate (sewage sludge single rate –SSs) contained 60 kg P per hectare and the double one (sewage sludge double rate –SSd) contained 120 kg P per hectare. The amount of organic matter introduced with sewage sludge with the single rate and double rate were as follows: in Estonia 1.59 and 3.18 t per hectare, in Germany 0.99 and 1.98 t per hectare, in Poland 1.97 and 3.94 t per hectare, respectively.

TABLE 1. GENERAL PROPERTIES OF THE SEWAGE SLUDGE APPLIED

Parameter	ESTONIA	GERMANY	POLAND
Dry matter [%]	29.79	23.30	43.96
Organic matter [% DM]	52.94	46.58	43.67
pH	7.88	n.d.[a]	11.05
Total content of macronutrients [g kg⁻¹ DM]			
N	22.7	23.1	27.5
P	20.0	28.2	13.3
K	3.0	n.d.	n.d.
Ca	55.0	n.d.	78.8
Mg	7.2	n.d.	8.8

[a]n.d. – not determined.

Soil bulk density (BD) and soil moisture content were determined in 100 cm³ of an undisturbed core sample after drying to constant weight at 105ºC. Soil specific density (SD) was measured by means of pycnometers. Soil total porosity (TP) was calculated, using the following equation (Kutílek and Nielsen 1994):

$$TP = \frac{(SD - BD)}{SD} \times 100 \quad [\%, v/v]$$

The porosity index was calculated according to the equation: porosity index = TP/(100-TP), where TP is total porosity.

Water permeability was determined in undisturbed soil samples taken to stainless steel sample rings (250 cm³). Soil samples were saturated in a water tank for 1–2 weeks and then placed in the laboratory permeameter. For the purpose of this study a permeameter type Eheim 1048 produced by the Eijkelkamp Agrisearch Equipment was used. Every sample was analysed in minimum ten replications. First the water flow in a unit of time (Q) was determined, then the coefficient K_t, at a laboratory temperature, and K_{10}, the coefficient of permeability at 10 °C were calculated.

For the purpose of the texture analysis, the pipette method, using the apparatus produced by Eijkelkamp Agrisearch Equipment (the International Standard –ISO 13317-2:2001) was applied. Soil material of ø< 2 mm was used after sample pre-treatment, i.e. organic matter removal by oxidation with 30% H_2O_2 and removal of calcium carbonates by HCl solutions. Soil texture classes were given in accordance with the Polish classification of the Polish Society of Soil Science [8].

RESULTS AND DISCUSSION

Estonian and German soils had sandy loam, and Polish soils had silt loam texture (Table 2). The highest content of clay fraction was found in Polish soils (19.26%) and the lowestone – in Estonian soils. The highest content of silt fraction was recorded in Polish soils (66.73%), twice lower content was found in Estonian ones (29.15%) and the lowest – in German ones (21.77%). Contrary to the silt content, German soils contained the highest amount (62.03%) and Polish soils the lowest amount (14.02%) of sand. Differences in grain-size-distribution within soils taken from various plots of the field trails were small. Taking into consideration the size of an individual plot (20 m^2), the differences in granulometric composition could be attributed to natural soil variability. In other words, the application of sewage sludge in the doses introduced did not change granulometric composition of the soils studied. However, prolonged application of sewage sludge could enhance the content of fine clay particles, as this organic fertiliser contains not only organic matter but also substantial amounts of colloidal mineral particles.

TABLE 2. TEXTURE OF STUDIED SOILS

	Plot	Percentage of particles (ø in mm)					Soil texture
		>2.0	2.0–0.05	0.05–0.02	0.02–0.002	<0.002	
MN	Estonia	3.99	58.53	10.93	16.41	14.13	sandy
(SD)	Control	(1.14)	(2.53)	(1.87)	(2.55)	(1.26)	loam
MN	Estonia	4.42	57.14	11.93	16.97	13.96	sandy
(SD)	SSs	(0.65)	(4.65)	(3.24)	(3.27)	(1.57)	loam
MN	Estonia	4.86	53.51	14.00	17.20	15.29	sandy
(SD)	SSd	(0.28)	(6.60)	(5.28)	(1.97)	(3.68)	loam
MN	Germany	9.98	60.75	7.16	15.13	16.96	sandy
(SD)	Control	(3.73)	(3.24)	(0.88)	(2.13)	(1.77)	loam
MN	Germany	8.69	63.53	7.67	13.81	15.00	sandy
(SD)	SSs	(2.02)	(4.16)	(2.21)	(1.53)	(2.20)	loam
MN	Germany	9.18	61.82	7.34	14.19	16.65	sandy
(SD)	SSd	(2.54)	(2.87)	(1.23)	(0.81)	(2.32)	loam
MN	Poland	0.96	13.65	34.81	31.85	19.69	silt
(SD)	Control	(0.60)	(1.66)	(2.64)	(2.39)	(2.29)	loam
MN	Poland	1.23	15.05	33.25	34.73	16.97	silt
(SD)	SSs	(0.39)	(2.30)	(2.95)	(2.32)	(1.90)	loam
MN	Poland	1.31	13.35	32.73	32.80	21.13	silt
(SD)	SSd	(0.13)	(2.32)	(1.09)	(2.08)	(3.86)	loam

MN – mean; SD – standard deviation.

    In Estonian test-field soils, sewage sludge application did not contribute to distinct changes in soil bulk density (BD) and soil specific density (Table 3). The analysed soils were appropriately compacted for agricultural use (BD 1.2 g cm^{-3}). Soil total porosity (TP) amounted to 53% in the plots without any treatment. After sewage sludge application, a small decrease in soil total porosity was observed (Table 3). In German soils, soil bulk density and soil total porosity increased after sewage sludge application but without correlation with the rate of sewage sludge applied. The soil porosity index was low and amounted to 0.5. Taking into consideration the bulk density of those soils, they could be classified as well-compacted. In Polish soils, similarly to Estonian plots, no distinct changes in soil bulk density (BD) and soil specific density were recorded. The analysed soils were compacted (BD 1.4 g cm^{-3}). Soil total porosity (TP) amounted to 41.7% in the plots without sewage sludge treatment and increased after sewage sludge application up to 43.32% in the plots where the double rate of sewage sludge had been applied (Table 3). Sewage sludge application modified total porosity, bulk density, and solid density of the test field soils. However, the changes were minor because the rates of SS were small. Statistical analysis for soil bulk density and soil total porosity did not show significant differences between control plots and plots fertilised with sewage sludge. Similar results were obtained by Łachacz *et al*. [5]. In order to significantly change soil bulk density and total porosity, higher rates of sewage sludge would have to be applied.

TABLE 3. PHYSICAL PROPERTIES OF STUDIED SOILS

	Plot	Bulk density of dry soil [g cm^{-3}]	Bulk density of moist soil [g cm^{-3}]	Specific density [g cm^{-3}]	Total porosity [%]	Porosity index*
MN	Estonia	1.235	1.537	2.600	53.19	1.106
(SD)	Control	(0.034)	(0.045)	(0.031)	(2.41)	(0.053)
MN	Estonia	1.243	1.578	2.608	52.34	1.099
(SD)	SSs	(0.036)	(0.040)	(0.041)	(1.18)	(0.052)
MN	Estonia	1.239	1.544	2.553	52.36	1.061
(SD)	SSd	(0.041)	(0.013)	(0.072)	(0.96)	(0.067)
MN	Germany	1.643	1.960	2.560	35.84	0.560
(SD)	Control	(0.070)	(0.069)	(0.038)	(2.46)	(0.059)
MN	Germany	1.681	2.005	2.534	33.65	0.509
(SD)	SSs	(0.045)	(0.034)	(0.055)	(2.82)	(0.065)
MN	Germany	1.668	1.986	2.558	34.76	0.534
(SD)	SSd	(0.025)	(0.025)	(0.052)	(2.02)	(0.046)
MN	Poland	1.452	1.655	2.493	41.70	0.717
(SD)	Control	(0.030)	(0.053)	(0.106)	(1.77)	(0.052)
MN	Poland	1.415	1.616	2.483	42.93	0.758
(SD)	SSs	(0.067)	(0.095)	(0.075)	(3.79)	(0.120)
MN	Poland	1.429	1.633	2.521	43.32	0.766
(SD)	SSd	(0.055)	(0.069)	(0.079)	(2.15)	(0.070)

* Porosity index = total porosity/(100-total porosity).

The water permeability coefficient (K_{10}) of Estonian soils varied greatly among plots and amounted from 0.834 m d^{-1} to 116.072 m d^{-1} (Table 4). Sewage sludge application increased water permeability of the soils studied. Water permeability of soil at Polish willow plots, similarly to Estonian ones, varied (Table 4). Many soils had low permeability coefficient (K_{10}) – below 1.0 m d^{-1}. Water permeability of the analysed German soils was generally low (Table 4). The water permeability coefficient showed the highest values in Estonian soils, which is connected with soil bulk density and total porosity. Bulk density of those soils was the lowest and total porosity was the highest among investigated soils. The lowest permeability was recorded in German soils and their bulk density was the highest.

TABLE 4. WATER PERMEABILITY OF STUDIED SOILS

	Plot	Q	K_t	K_{10}
		[cm³ min.$^{-1}$]	[m d^{-1}]	[m d^{-1}]
MN	Estonia Control	2.476	18.190	22.023
SD		1.974	10.392	12.598
ME		2.593	16.978	20.543
MN	Estonia SSs	2.728	18.624	22.576
SD		3.613	17.916	21.638
ME		1.310	17.822	21.647
MN	Estonia SSd	4.037	20.993	25.491
SD		3.850	31.839	38.490
ME		3.003	9.458	11.599
MN	Germany Control	0.086	0.365	0.453
SD		0.153	0.775	0.961
ME		0.032	0.026	0.033
MN	Germany SSs	0.117	0.386	0.481
SD		0.139	0.763	0.946
ME		0.074	0.061	0.078
MN	Germany SSd	0.031	0.042	0.054
SD		0.053	0.070	0.089
ME		0.008	0.016	0.021
MN	Poland Control	2.604	4.294	5.375
SD		4.234	7.825	9.810
ME		0.585	0.729	0.905
MN	Poland SSs	0.583	0.552	0.695
SD		0.590	0.573	0.729
ME		0.516	0.443	0.551
MN	Poland SSd	3.833	5.756	7.195
SD		4.502	7.617	9.478
ME		1.451	1.976	2.459

ME – median; Q – volume of water flowing through the sample during a period of time; K_t – permeability coefficient at the applied temperature; K_{10} – permeability coefficient at 10 ºC (corrected coefficient at 10 ºC).

According to permeability classes for agriculture and conservation elaborated by the FAO [13], Estonian soils have very rapid, moderate, and moderately rapid class of water permeability. German control soils have moderately slow permeability class. The increase in permeability was recorded for soils fertilised with the single rate of SS (moderate permeability) whereas soils fertilised with the double rate of SS had very slow permeability class. In German soils no relation between the rate of SS and permeability was stated. Polish control soils show rapid permeability. However, the differences between the examined soils were considerable (K_{10} was from 0.025 to 27.847 m d^{-1}). Contrary to German tested soils, the decrease in water permeability was recorded for Polish soils fertilised with the single rate of SS (moderate permeability) and the increase – for soils fertilised with the double rate of SS (very rapid permeability).

Sewage sludge application as well as the way of land use influenced water permeability. The presence of small stones, earthworms or sprouts in agriculturally used soils influences water permeability. The volume of pores greater than 150 micrometers also increases after application of organic amendments, which has an impact on water permeability in soil [6, 7].

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PRZEPUSZCZALNOŚĆ WODNA GLEB NAWOŻONYCH OSADEM ŚCIEKOWYM NA PLANTACJACH W EUROPIE

Przepuszczalność wodna gleb charakteryzuje przemieszczanie się wody w profilu glebowym. Zależy przede wszystkim od wielkości i geometrii porów glebowych, gęstości objętościowej i struktury gleby. Badane gleby na plantacjach wierzby nawożone osadem ściekowym w Polsce, Estonii i Niemczech. Zastosowanie osadu ściekowego oraz użytkowanie gleby miało wpływ na przepuszczalność wodną badanych gleb.