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FRESHWATER MUSSEL SPECIES
FUSCONAIA ASKEWI, FUSCONAIA
LANANENSIS, AND PLEUROBEMA
RIDDELLII

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POPULATION DYNAMICS AND GENETIC DIFFERENTIATION OF THE
THREATENED FRESHWATER MUSSEL SPECIES
FUSCONAIA ASKEWI, *FUSCONAIA LANANENSIS*, AND *PLEUROBEMA RIDDELLII*

by

EDITH PLANTS-PARIS

A thesis submitted in partial fulfillment
of the requirements for the degree of
Master of Science
Department of Biology

Neil B. Ford, PhD., Committee Chair
College of Arts and Sciences

The University of Texas at Tyler
May 2016

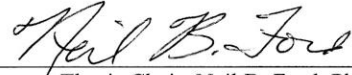
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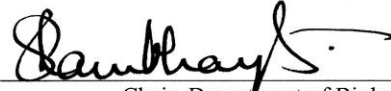
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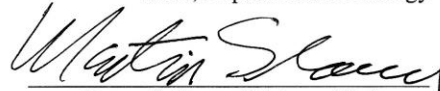
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Abstract

POPULATION DYNAMICS AND GENETIC DIFFERENTIATION OF THE THREATENED FRESHWATER MUSSEL SPECIES *FUSCONAIA ASKEWI*, *FUSCONAIA LANANENSIS*, AND *PLEUROBEMA RIDDELLII*

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North America has the most diverse freshwater mussel fauna in the world with approximately 300 species; unfortunately, extinction rates for freshwater mussels rivals the rates of many other groups of organisms. Population-level natural life history data is essential in the management of species of conservation concern, yet basic information about freshwater mussel life-history and demographic traits are unknown for many species. To further complicate matters, taxonomic uncertainty exists among some members of the group. The work detailed herein had two goals: to gain further understanding of the taxonomic relationship between *Fusconaia lananensis* and *F. askewi* by sequencing genes that had not been previously examined for these species, genes *16S* and *ITS1*, and collect data on the population size, density, and structure for both *F. askewi* and *F. lananensis*, as well as for *Pleurobema riddellii*, all of which are classified as state threatened in Texas. The second goal was accomplished via qualitative analysis of data from 0.25 m² quadrats and through mark-recapture studies at field sites where the highest densities of these species have been recorded. Specifically, quadrat

surveys were conducted at seven mark-recapture sites in the Neches, Sabine, and Angelina Rivers during the summers of 2014 and 2015. In terms of my genetic analysis, data collected from the *16S* gene has provided additional support that *F. askewi* and *F. lananensis* are one single species, as recently proposed by other researchers. Data collected from the *ITS1* gene showed no genetic differentiation between *F. askewi*, *F. lananensis*, and *F. flava*, though recently published research indicates that there is low genetic variation within the *ITS1* gene for several different species found in genus *Fusconaia*. Sites on the Neches and Angelina Rivers had significantly higher recapture rates between 2014 and 2015 than sites on the Sabine River, likely because of a flooding event that occurred in the Sabine River during that time. The largest population estimate for a *F. askewi* population was 302 ± 26.72 in 2015 within the 25 m area while the largest population estimate for a *P. riddellii* population was 101 ± 4.99 in 2015 within the 25 m area. Fewer juvenile *P. riddellii* were detected than *F. askewi*, leading to left-skewed size class distributions for *P. riddellii*. As conservation efforts for freshwater mussels increase, continued analysis of established freshwater mussel populations will be crucial.

Chapter 1: Introduction to Freshwater Mussels

North America has the most diverse freshwater mussel fauna in the world with approximately 300 species (Williams et al. 1993). Often a large percentage of the benthic biomass in freshwater systems is comprised of freshwater mussels, and thus mussels likely play an integral role in these systems. Specifically, through their burrowing and filtration feeding behaviors freshwater mussels likely are critical in nutrient cycling (Vaughn and Hakenkamp 2001). At the same time, freshwater mussels are an important food source for other organisms at higher trophic levels such as mammals, birds, and fishes (Haag 2012). Freshwater mussels and their shells also play a role in substrate stability and in habitat heterogeneity (Gutiérrez et al. 2003).

Historically freshwater mussels have been recognized as efficient suspension feeders that consume primarily plankton (Dame et al. 1985, Dame et al. 1991). Freshwater mussels have been observed filtering large amounts of water within short periods of time (Kryger and Riisgard 1988, McIvor 2004, Strayer et al. 2004). Along with plankton, dissections of mussel digestive tracts have also contained zooplankton and detritus. Freshwater mussels have also been documented feeding off of sediment through the use of their foot and through use of their siphon on the sediment surface (Strayer et al. 2004, Nichols et al. 2005). Feeding methods employed by freshwater mussels have been shown to affect nutrient cycling in their ecosystem in several ways such as transferring nutrients from the water column to the riverbed and stimulating both primary and secondary production through nutrient excretion (Spooner and Vaughn 2006, Spooner 2007, Vaughn et al. 2008).

Unlike most marine mussels, freshwater mussels have specialized larvae, called glochidium, which parasitically feed from the gills of host fish (Strayer et al. 2004). Techniques and adaptations to transmit glochidia to hosts vary between freshwater mussel species. One specialized method described is the use of lures by the females of some species to mimic fish or invertebrates (Haag and Warren 2000, Haag and Warren 2003). In other species, female mussels release their glochidia in conglomerates that mimic fish food sources such as eggs or larvae and thus are attacked by host fish (Jones and Neves 2002, Haag and Warren 2003). These two methods typically attract a few specific fish species and are seen within mussel species that are considered host specialists. A third method of glochidia transmission is used by mussels considered host generalists in which a mucous web of glochidia that entangle fish indiscriminately is released (Haag and Warren 2003). These adaptations in glochidia transmission to fish-hosts indicate a close evolutionary link between mussel life-history traits and their use of host-fish (Strayer et al. 2004).

Unfortunately, extinction rates for freshwater mussels rival the rates of many other groups of organisms (Ricciardi and Rasmussen 1999). As of 2014, 28% of freshwater mussels were federally listed as imperiled but some researchers suggest that this number could be as high as 65% (Haag and Williams 2014). There are multiple factors that may impact freshwater mussel fauna, but the destruction of river systems by the creation and use of dams may cause the most impact with both positive and negative consequences on mussel populations (Singer and Gangloff 2011, Haag 2012, Gangloff 2013). In the last decade or so, mussel conservation efforts have greatly increased via federal and state agencies in the United States, as well as various other conservation

groups, though freshwater mussel populations still face threats of decline (Haag 2012). Knowledge of the various aspects of life-history traits is crucial for protecting freshwater mussels; unfortunately, much of this information is unknown for many species of freshwater mussels.

Freshwater Mussels of East Texas

52 species of freshwater mussels in the family unionidae are found in Texas and 15 are currently listed as state threatened by the Texas Parks and Wildlife Department. Of these 15 species, 6 are found in the East Texas area including: Southern Hickorynut (*Obovaria jacksoniana*), Texas Heelsplitter (*Potamilus amphichaenus*), Sandbank Pocketbook (*Lampsilis satura*), Louisiana Pigtoe (*Pleurobema riddelli*), Texas Pigtoe (*Fusconaia askewi*), and the Triangle Pigtoe (*F. lananensis*). Despite being of conservation concern, little is known about the various aspects of these species' population biology, genetics, and their life history traits (Howells, unpublished). Though general ranges are known for these species, there are few sites that are known to contain a high abundance of any of these threatened species.

The species targeted in the current study were the 3 state-threatened pigtoe species *P. riddelli*, *F. askewi*, and *F. lananensis*. Though there has been taxonomic uncertainty between *F. lananensis* and *F. askewi*, *P. riddelli* is morphologically and genetically distinct from *Fusconaia* species (Burlakova et al. 2012, Howells et al. 2012). While previous studies have gathered additional information about these three species, including the creation of habitat suitability models (Ford 2013) and the identification of a fish host for *F. askewi* (Marshall 2014, Bertram 2015), much remains to be learned about these three species. The goal of this project was to 1) to gain further understanding of the

taxonomic relatedness of *F. lananensis* and *F. askewi* by sequencing genes that have not been previously sequenced for these species (*16S* and *ITS1*), and 2) to study population size, density, and structure for *P. riddellii*, *F. askewi*, and *F. lananensis* through quantitative analysis of data from the use of 0.25 m² quadrats and through mark-recapture studies at high-density field sites.

Chapter 2: Genetic Differentiation of *Fusconaia askewi* and *F. lananensis*

Traditionally, species descriptions for freshwater mussels were based largely on shell morphology, which can vary greatly between individuals in a population and along environmental gradients (Haag 2012). This creates difficulties in defining species based solely on morphology and has led to researchers incorporating molecular genetic data to help define and identify species, with the mitochondrial genes *NDI* and *COXI*, and the nuclear genes *ITS1* and *ITS2* most commonly used for freshwater mussels (e.g. Burlakova et al. 2012, Kallersjo et al. 2005, Inoue et al. 2014). Although mitochondrial genes are commonly used when conducting molecular phylogenetic analyses with freshwater mussels, species found within bivalve families Unionidae, Veneridae, and Mytilidae are known to use a unique method of mitochondrial inheritance that makes them less useful. Specifically, female mussels transmit their mitochondria to all of their offspring (F-type), but male mussels can also transfer their mitochondria (M-Type) to their sons resulting in heteroplasmic males. Some studies have sequenced both mitochondrial types but the effect of heteroplasmy on conclusions from genetic analyses is not understood. Therefore nuclear genes such as *ITS1* and *ITS2* in addition to mitochondrial genes should be used in studies involving freshwater mussels (Krebs 2004, Mock et al. 2004, Kallersjo et al. 2005).

F. lananensis (Triangle Pigtoe) and *F. askewi* (Texas Pigtoe) are difficult to distinguish and currently both are listed as State threatened by the Texas Parks and Wildlife Department. Since, *F. lananensis* and *F. askewi* are often sympatric, proper identification of the two is critical. *Fusconaia askewi* occurs from the San Jacinto River north to the Red River system and *F. lananensis* is found in the Angelina River, the

Neches River, and the Attoyac River (Howells 2014). Along with range overlap, *F. lananensis* and the *F. askewi* also possess morphological similarities in external and internal characteristics. Both *F. lananensis* and the *F. askewi* may possess a sub-rectangular shape and similar external coloration ranging from chestnut brown to black (Howells et al. 2012, Howells 2014). Because of their similar external morphology and with additional variation existing within each species, these species typically cannot be positively identified in the field (Figure 2.1, Howells 2014). Additional variation can typically be seen inside the shell for both species, particularly with coloration of the nacre. *Fusconaia askewi* is often characterized as having white nacre with pink or red nacre outside the palatal line and *F. lananensis* is often identified as having solid pink nacre with occasional pearly bumps and yellow blotches (Figure 2.1, Howells 2014). *F. askewi* has also been observed having solid pink, orange or solid white nacre. Given that such variation exists in the nacre, using coloration may not be a reliable method for species identification if the distinguishing pearly bumps and yellow blotches are not seen for *F. lananensis* or the white and pink coloration of the nacre in *F. askewi*. Both species are described as having three pseudocardinal teeth (two left, one right) triangular and compressed and three lateral teeth (two left, one right) straight to slightly curved (Howell 2014).



Figure 2.1. Visual comparison of *Fusconaia lananensis* (left) and *F. askewi* (right) external and internal morphology. Both specimens were collected in the Angelina River off 343.

Similarities in morphology and geographic ranges lead Burlakova et al. (2012) to hypothesize that *F. lananensis* and *F. askewi* are not separate species. She found low genetic variation within the *NDI* and *COXI* genes between *F. askewi* and *F. lananensis* (1% for *NDI* and 0.7% for *COXI*) (Burlakova et al. 2012). However, other authors expressed concern over the methodology in this study, including: 1) the full range of morphological variation between *F. lananensis* and *F. askewi* were not acknowledged in this study and 2) collection sites for mussels were not identified (Howells et al. 2012). However, other recent studies using the *NDI* and *COXI* genes also support the idea that *F. askewi* and *F. lananensis* are not separate species (Marshall 2014, Bertram 2015). While both species are currently listed as threatened in the state of Texas because of the

rarity of *F. lananensis*, the difficulty distinguishing *F. askewi* from *F. lananensis* has inhibited determination of whether *F. askewi* might actually not be in need of protection. The goal of this study was to include genes *16S* and *ITS1* to the analysis of the genetic relationship between *F. askewi* and *F. lananensis* to add more genetic data to this analysis. Genes *ND1* and *COXI*, which have been used in earlier studies with these species, were also included in the analysis as additional localities for specimens were sampled. All study site locations were recorded for all sequenced individuals and a larger sample size of *F. lananensis* was included in this study. All specimens were collected and preserved to help confirm species identification with the addition of internal morphological characteristics.

Materials and Methods

I collected *F. lananensis* and *F. askewi* in the summers of 2014 and 2015 from the Neches and Angelina Rivers (Table 2.1). Whole specimens were frozen so that the shells could be referred to during later analyses and to confirm species identification through internal morphology. Foot and mantle tissue was removed and preserved in 95% ethanol and stored at -20°C for DNA analysis. DNA from preserved tissue samples were extracted using E.Z.N.A. Gel Extraction Kits (Omega bio-tek, Norcross, GA). The mitochondrial genes (mtDNA) *ND1*, *COXI*, *16S* and nuclear gene *ITS1* were then amplified via polymerase chain reactions (PCRs) for each individual. Primers used were as follows:

ND1: 5' -TG GCAGAAAAGTGCATCAGATTAAGC-3'

5' -TCGGAATTCTCCTTCTGCAAAGTC-3'

(Serb et al. 2003)

COX1: 5' -GTTCCACAAATCATAAGGATATTGG-3'

5' -TACACCTCAGGGTGACCAAA AAACCA-3'

(Campbell et al. 2005)

16S: 5' -CCGTTCTGAACTCAGCTCATGT-3'

5' -CGACTGTTTAACAAAAACAT-3'

(Campbell et al. 2005)

ITS1: 5'-AAAAAGCTTCCGTAGGTGAACCTGCG-3'

5'-AGCTTGCTGCGTTCTTCATCG-3'

(King et al. 1999)

PCR parameters for *ND1*, *COX1*, and *ITS1* were as follows: 94° C for 5 m, 30 cycles of 94° C for 45 s, 54° C for 60 s, and 72° C for 60 s followed by a final extension of 72° C for 5 m (King et al. 1999, Campbell et al. 2005, Serb et al. 2003). PCR parameters for *16S* were as follows: 92° C for 5 m; 92° C for 40 s, 50° C for 60 s, 68° C for 90 s, x 35; 72° C for 10 m (Campbell et al. 2005). An Eppendorf Mastercycler gradient thermal cycler was used to amplify all PCR reactions. Gel electrophoresis was used to test the quality of amplification and successfully amplified PCR products were purified using E.Z.N.A. cycle pure kits (Omega bio-tek, Norcross, GA) following the standard protocol with an additional 30 µL of purified water for resuspension. Purified DNA was concentrated to 17-20 ng/ µL with a 260/280 ratio around 1.8 to 2.0 as recommended by Eurofins MWG Operon where reactions were shipped for sequencing using BigDye Terminator v 3.1 Cycle Sequencing kits (Applied Biosystems).

Table 2.1 Tissue samples from *Fusconaia lananensis* and *F. askewi* used in the final DNA analysis

Species	Sample #	Location Collected	Genes used in analysis
<i>F. askewi</i>	EP130	Neches, Downstream 294	<i>NDI, ITS1</i>
<i>F. askewi</i>	EP134	Neches, Downstream 294	<i>NDI, ITS1</i>
<i>F. askewi</i>	EP136	Neches, Downstream 294	<i>NDI, 16S, ITS1</i>
<i>F. askewi</i>	EP138	Neches, Downstream 294	<i>NDI, 16S</i>
<i>F. askewi</i>	EP139	Neches, Downstream 294	<i>NDI, 16S, ITS1</i>
<i>F. lananensis</i>	EP106	Angelina, Downstream 343	<i>COXI, ITS1</i>
<i>F. lananensis</i>	EP107	Neches, of 79	<i>NDI, COXI, ITS1</i>
<i>F. lananensis</i>	EP145	Angelina, Upstream 343	<i>NDI, COXI, 16S, ITS1</i>
<i>F. lananensis</i>	EP146	Angelina, Upstream 343	<i>NDI, COXI, ITS1</i>
<i>F. lananensis</i>	EP150	Angelina, Upstream 343	<i>NDI, COXI, 16S, ITS1</i>
<i>F. lananensis</i>	EP151	Angelina, Upstream 343	<i>NDI, COXI, 16S</i>
<i>F. lananensis</i>	EP152	Angelina, Upstream 343	<i>NDI, 16S, ITS1</i>
<i>F. lananensis</i>	EP153	Angelina, Upstream 343	<i>NDI, 16S, ITS1</i>
GenBank Sequences			
<i>F. flava</i>	AY613793	Campbell et al. 2005	<i>NDI</i>
<i>F. askewi</i>	JN180998	Burlakova et al. 2012	<i>COXI</i>
<i>F. askewi</i>	KT285626	Pfeiffer et al. 2015	<i>COXI</i>
<i>F. flava</i>	KT285636	Pfeiffer et al. 2015	<i>COXI</i>
<i>F. flava</i>	AY238481	Krebs et al. 2003	<i>16S</i>
<i>F. flava</i>	DQ383442	Campbell et al. 2008	<i>ITS</i>

After the DNA sequences were obtained, the sequences were edited and aligned using programs Sequencher, Clustal X, and Mesquite (Gene Codes 2000, Larkin et al. 2007, Maddison and Maddison 2004). Program AliView was used alongside Mesquite for visualizing the DNA sequences (Larson 2014). Sequences from GenBank (<http://www.ncbi.nlm.nih.gov>) were used to compare to out putative *F. lananensis* and *F. askewi* sequences to related species *Fusconaia flava*, the Wabash Pigtoe, and to provide additional *F. askewi* sequences for analysis of the *COXI* gene (additional *COXI* sequences for *F. lananensis* were not available via GenBank). Percent divergence values

were calculated for all genes, comparing 1) individuals identified as *F. lananensis* and *F. askewi* based on morphology and collection locale and 2) *F. lananensis* and *F. askewi* to *F. flava* to compare divergence values with a species that is genetically distinct from the study species.

Results

The *NDI* gene was successfully sequenced for a total of 14 samples (Table 2.2). Sequencing was less successful with the *COXI* gene only sequenced for 7 individuals, the *16S* gene sequenced for 10 individuals, and the *ITS1* gene sequenced for 12 individuals (Table 2.2). The *COXI* gene could not be sequenced from the *F. askewi* samples. For this gene, two *F. askewi COXI* sequences were obtained via GenBank and compared to the *F. lananensis* sequences for analysis. *NDI* sequences were trimmed to a length of 764 bases long and were compared to a *F. flava* sequence from GenBank (Table 2.1). When comparing the *F. lananensis* and *F. askewi NDI* sequences to the *F. flava NDI* sequence, the sequences differed by 3.14% (Table 2.2). When comparing *F. lananensis* sequences to *F. askewi* sequences, the percent divergence dropped to 0.39% (Table 2.2) *F. askewi* and *F. lananensis* both had percent divergence values of 0.26% when looking at variation within each species (Table 2.3). *COXI* sequences were trimmed to a length of 604 bases long. When compared to the *F. flava* sequence a percent divergence of 4.14% was calculated (Table 2.2). When removing the *F. flava* sequence *F. lananensis* and *F. askewi* were 0.99% divergent (Table 2.2). Looking at the variation with the two species, *F. lananensis* had a slightly higher percent divergence value than *F. askewi* within the *COXI* gene (Table 2.3). The *16S* sequences were trimmed to a length of 444 bases long. When compared to *F. flava* a percent difference of 5.63% was calculated for all

sequences (Table 2.2). When excluding *F. flava* a percent difference value of 1.351% was calculated for *F. lananensis* and *F. askewi* (Table 2.2). Most of this variation came from a single sequence, EP151; when this sequence was removed from the analysis the percent difference dropped to 0.23%. For gene *16S* no individuals shared mutations; all observed mutations were unique. Looking at the variation within each species, there was no divergence within the *F. askewi* samples for the 16S gene (Table 2.3). Gene *ITS1* was trimmed to a length of 507 bases long. A percent divergence value of 0.80% was calculated for analysis with and without *F. flava* (Table 2.2). Comparing the variation within each species, *F. lananensis* had a higher percent divergence value than *F. askewi*, with *F. lananensis* having a value of 0.39% and *F. askewi* having a value of 0.79% (Table 2.3).

Table 2.2. Percent divergence values for *Fusconaia lananensis* and *F. askewi*, with the number of individuals used in the final analysis for each gene is included. All *F. flava* sequences used in the analyses were obtained through GenBank.

	<i>NDI</i>	<i>COX1</i>	<i>16S</i>	<i>ITS1</i>
Number of <i>Fusconaia askewi</i>	5	2*	4	5
Number of <i>Fusconaia lananensis</i>	7	6	5	7
Percent Divergence				
<i>Fusconaia askewi/Fusconaia lananensis</i>	0.39	0.99	1.35	0.80
<i>Fusconaia askewi/Fusconaia lananensis</i> + <i>Fusconaia flava</i>	3.14	4.14	5.63	0.80

* Indicates samples obtained from GenBank for use in the analysis

Table 2.3 Percentage divergence values for *F. lananensis* and *F. askewi*, with divergence values within species.

	<i>NDI</i>	<i>COX1</i>	<i>16S</i>	<i>ITS1</i>
<i>Fusconaia askewi</i>	0.26	0.50	0	0.39
<i>Fusconaia lananensis</i>	0.26	0.66	1.35	0.79
<i>Fusconaia askewi/Fusconaia lananensis</i>	0.39	0.99	1.35	0.80

Discussion

Percent divergence values obtained from this study for the *NDI* and *COXI* genes are similar to values found in other genetic studies involving *F. lananensis* and *F. askewi*; specifically, a divergence value of 0.39% was calculated for the *NDI* gene when comparing *F. lananensis* with *F. askewi* in the current study and this is congruent with the values arrived at by Burlakova et al. (2012) and Marshall (2014). When *F. flava* is included in the analysis I calculated a value of 3.14% while Burlakova et al. (2012) calculated a range from 2.59-3.43%. Divergence values calculated for *COXI* were slightly higher than values obtained in a previous study; specifically, a divergence value of 0.99% was calculated for *F. lananensis* and *F. askewi* individuals in my study while values ranging from 0.3-0.7% were seen in Burlakova et al. (2012). Divergence values between *F. lananensis* and *F. askewi* with the addition of *F. flava* for the *COXI* gene were similar to values obtained with Burlakova et al. (2012).; a value of 4.14% was calculated in the current study when including *F. flava* while Burlakova et al. (2012) had a range of 2.92-4.91%. Studies looking at percent divergence for the genus *Fusconaia* for the *I6S* have not been conducted prior to the current study; however, the amount of divergence found between *F. flava* and *F. lananensis*/*F. askewi* for this study was 5.63%, similar to the amount of variation seen within different species of the same genus in other studies (Kallersjo et al. 2005). The amount of variation within just *F. lananensis*/*F. askewi* dropped to a much lower value of 1.35% without the presence of *F. flava*. There was no difference in divergence values between the *F. lananensis* and *F. askewi* samples and the *F. flava* sample (Table 2.2). This is similar to results for individuals of different species in *Fusconaia* that show little to no variation with the *ITS1* gene (Manendo et al.

2008, Schilling 2015). This could indicate that these species within *Fusconaia* are more recently diverged; yet there is also evidence that variation within the *ITS* gene regions could be species specific (Kallersjo et al. 2005, Manendo et al. 2008). There are three regions within the *ITS* gene that can be used for sequencing: the *ITS1*, 5.8s, and the *ITS2* region. The *ITS1* and the *ITS2* regions can have a large amount of diversity based on species, with values of <1% reported within species and 4.7%-15.3% between species (Kallersjo et al. 2005). As the *ITS1* gene was used in this analysis it is possible that the *ITS2* region could present more variation for *Fusconaia*.

Overall, the percent divergence values based on the amount of variation within the *ND1*, *COX1*, and *16S* genes from the *F. lananensis* and *F. askewi* do not support the current classification of these individuals as separate species and validates the work of Burlakova et al. (2012). Divergence values obtained with the *ND1* and *COX1* genes were similar to those found with Burlakova et al. (2012) both within the *F. lananensis* and *F. askewi* and with the inclusion of *F. flava*. The addition of the results for the *16S* gene further support the hypothesis that *F. lananensis* and *F. askewi* are not separate species. Though the *ITS1* gene did not confirm nor deny this, further analysis using different regions of the *ITS* gene could bring further support for combining the two species.

Chapter 3: Population Dynamic of Freshwater Mussels

The population structure and rate of growth for individuals in freshwater mussel populations can be determined through life history traits such as individual growth rates, life spans, and host interactions (Haag 2012). For example, traits related to the local population sizes of mussels in the Red River basin, such as regional abundance and time spend brooding, were strong predictors of local extinction (Vaughn 2012). Determining life history traits for freshwater mussel species can also help to understand species distribution in rivers (Haag and Warren 1998, Daniel and Brown 2014). In addition, freshwater mussel species dependent on host fish density are often restricted to sites with stable host populations (Haag and Warren 1998). Unfortunately, population structures and individual growth rates are not known for most species of freshwater mussels and generalizations are made about life history only based on well-studied species such as *Margaritifera margaritifera*, the Freshwater Pearl Mussel, found throughout Europe and parts of eastern Canada (Hastie et al. 2000, Outeiro et al. 2007). For example, depictions of all freshwater mussel species as being slow growing and long lived are not accurate. Though many species of mussels do have long lifespans, lifespans between freshwater mussels species vary from 4 to 200 years (Haag and Rypel 2011). As other species of freshwater mussels have been studied they have shown great differences in life-history traits, the need to study these traits in individual species has become more apparent (Haag 2012).

Age and size structure of populations have been studied in some species of freshwater mussels (Bauer 1983, Hastie et al. 2000, Rogers et al. 2001, Haag 2012) and

left-skewed size-class distributions are most commonly found (Rogers et al. 2001, Haag 2012). This pattern has been observed with species known to have low recruitment rates and high survival, but this left-skewed distribution may be caused by two other factors: 1) human impacts may be suppressing recruitment rates of some species and 2) surveys used with freshwater mussels tend to be biased against smaller individuals (Bauer 1983, Hastie et al. 2000, Haag 2012). To obtain unbiased information about a population, intensive excavation methods have been used to increase detectability of smaller individuals (Miller and Payne 1988, Haag and Warren 2007, Haag 2012). Using these methods, three types of size distributions which differ from traditional left-skewed distributions can be seen in a healthy stream: 1) cohort-dominated, 2) uniform, and 3) right-skewed distributions (Haag 2012). Cohort-dominated freshwater mussel populations are dominated by one or few size classes representing size/age cohorts with other classes represent by few individuals (Payne and Miller 1989, Payne and Miller 2000, Haag 2012). Populations with uniform distributions have a relatively even frequency of individuals across size classes and may be dominated by mid-sized or large individuals (Haag 2012). This occurs due to of the accumulation of older individuals as growth slows and is often seen with longer-lived species (Miller and Payne 1993, Haag and Warren 2007, Haag and Warren 2010, Haag 2012). Finally, right-skewed populations consist of classes dominated by younger individuals with smaller numbers of older individuals and are often seen with short-lived species (Crabtree and Smith 2009, Haag and Warren 2010, Jones and Neves 2011, Haag 2012).

Several different methods have been employed to determine different population characteristics, such as density of a species. The two most common methods of mussel

surveying include 1) quadrat sampling and 2) timed surveys (Vaughn et al. 1997).

Quadrat surveys are sometime also used in conjunction with timed surveys, as there are observed benefits and disadvantages to both methods. For example, timed surveys tend to underestimate small species and quadrat surveys tend to underestimated species richness in an area (Vaughn et al. 1997). Both 0.25 m² and 1 m² quadrats are commonly use to determining specific population characteristics such as species abundance, population density, distribution, and sizes (Vaughn et al. 1997, Kuenzler 2003, Strayer and Smith 2003).

Mark-recapture is another method used to study freshwater mussel populations although some concerns have been raised over potential bias towards larger individuals within a mussel population (Anthony et al. 2001, Haag 2009, Hua et al. 2015).

Depending on the models used, mark-recapture studies can be used to calculate several different aspects in a population including survival, recruitment, and population size (Matter et al. 2013). In addition, measurements of individual mussels can be tracked over a period of months or years to determine both individual and class growth rates (Villella et al. 2004). Results obtained from mark-recapture can be further analyzed to examine relationships between aspects such as population density and population growth with external influences (Lauzon-Guay et al. 2005, Widarto 2007). Recently mark-recapture studies have even been used in the field of freshwater mussel propagation to assess both the growth of lab-reared mussels in the wild and impacts they have when released on established populations (Hua et al. 2015). Several different methods of tagging have been used in mussel mark-recapture studies including glued tags, carved numbers on shells, and passive integrated transponder (PIT) tags (Peterson et al. 2011, Kurth et al. 2007,

Hua et al. 2015). Though animals marked with PIT tags have a much higher recapture rate than the use of glued tags and carved tags, glued tags and carved tags are currently cheaper to use (Kurth et al. 2007, Hua et al. 2015).

The goal of this project was to estimate population size, density, and structure for the state-threatened species *Pleurobema riddellii*, *Fusconaia askewi*, and *F. lananensis*. This was done by using 0.25 m² quadrats and mark-recapture at field sites where the highest densities of these three species have been recorded. Multiple mark-recapture sites were set up for each species to monitor population size and growth over the course of a year. As information on the population ecology for these threatened species is limited, additional data about these populations are crucial for conservation efforts.

Materials and Methods

Field Sites

Locations for the field sites used were selected from sites with the top highest densities for the target species in previous surveys (Ford 2013). From those locations, I selected sites that had easy accessibility via boat ramps as it was necessary to visit several times during the course of this study. During the summer of 2014, I established a total of seven field sites with three in the Neches River, three in the Sabine River, and one in the Angelina River (Table 1). Populations of *P. riddellii* were examined at the sites in the Neches River (Table 1). Populations of *F. askewi* were studied at sites on the Sabine River (Table 1). Finally, I established one site on the Angelina River to study a population consisting of both *F. askewi* and *F. lananensis* (Table 1). For this study, all

individuals under study were classified as *Fusconaia* because of the difficulty in differentiation between *F. askewi* and *F. lananensis* in the field.

Random 0.25 m² surveys

During the summer of 2014 and 2015 I conducted random 0.25 m² surveys at each of the seven field sites. Specifically, a 150-m segment of the river was marked off at the field site and divided into three 50-m segments. Within each 50-m segment, 27 0.25 m² quadrats were sampled using a stratified randomize design with three starts: one near each bank and one in the middle of the river (Strayer and Smith 2003, Pooler and Smith 2005). All mussels within each quadrat were excavated by hand by student workers. I recorded all live and recently dead mussel species and measured length, height, and width of the 3 study species. I used the results from these surveys to help choose where the 5 m x 5 m mark-recapture site would be set up for the mark-recapture study. The 0.25 m² quadrat with the largest number of the desired species was the point where the 25 square meter mark recapture location was established. Therefore those sites had the highest density of mussels. Most of the 0.25 m² surveys were conducted during the summer of 2014, except for sites Sabine 2 and Sabine 3. The 0.25 m² survey data for these sites for the summer of 2015 were obtained from surveys conducted by another graduate student (Jared Dickson, unpublished).

Table 3.1. Site names, locations, coordinates, and species of study for each field site.

<i>Site Name</i>	<i>River</i>	<i>Location</i>	<i>Coordinates</i>	<i>Species</i>
<i>Neches 1</i>	Neches	Upstream Hwy 294	N 31.643610, W-95.285900	<i>Pleurobema riddellii</i>
<i>Neches 2</i>	Neches	Cherokee Hunting Club	N 31.715680, W-95.332570	<i>P. riddellii</i>
<i>Neches 3</i>	Neches	Downstream Hwy 79	N 31.841370, W-94.425150	<i>P. riddellii</i>
<i>Sabine 1</i>	Sabine	Downstream Hwy 14	N 32.553450, W-95.200690	<i>Fusconaia askewi</i>
<i>Sabine 2</i>	Sabine	Hwy 14 Bridge	N 32.557638, W-95.205906	<i>F. askewi</i>
<i>Sabine 3</i>	Sabine	Upstream Hwy 43	N 32.377156, W-94.465937	<i>F. askewi</i>
<i>Angelina 1</i>	Angelina	Upstream 343	N 31.753400, W-94.961610	<i>Fusconaia spp.</i>

Mark-Recapture

A total of seven sites were chosen for mark and recapture research: three on the Sabine River, three on the Neches River, and one on the Angelina River. As indicated above I chose these sites as they had the highest density locations for the targeted species. The sites on the Sabine were chosen for *F. askewi* mark/recapture, the sites on the Neches were chosen for *P. riddellii* mark/recapture, and the site on the Angelina was chosen for *F. lananensis* and *F. askewi* mark/recapture. Each 5 m x 5 m area was bounded by rebar rod positioned in the banks of the river, and by spray painting trees on the bank, and by taking a GPS point within the middle of the mark-recapture location. We placed the 1 m² quadrats within the area and excavated all of mussels by hand. After we completed excavating one the quadrat, we moved the quadrat and excavated the next 1 m² area. We excavated the entire 5 m x 5 m area for a total of 25 quadrats. In instances where the river was too deep for excavation without diving or unsuitable habitat for mussels would have fallen within the 5 m x 5 m square, a different shape equaling approximately 25 m² was

used instead. I marked all live *P. riddellii*, *F. askewi*, and *F. lananensis* mussels by gluing bee queen-marking tags on the mussel shell with super glue. The tag color, tag number, mussel species, and mussel length were recorded for each mussel marked. After marking, each mussel was carefully placed back within the quadrat.

All mark-recapture sites were visited an additional two times during the summer of 2015 for the 1-year mark-recapture period and for the 2-3 week mark-recapture period. During the first visit during the summer of 2015, we re-excavated the sites for mussels through use of the 1 m² quadrats, including the immediate area outside the 5 m x 5 m quadrat. I recorded the tag color, tag number, species, and length for all previously marked mussels found. I also recorded the number of marked and unmarked target species mussels and tagged and recorded all new mussels found during that time. For the field sites in the Neches River and the Angelina River, we tagged all mussels with the Biomark PIT tag. PIT tags were attached to the outer shell with superglue, covered with a marine epoxy, and then allowed to dry for approximately 15 minutes. I recorded the PIT tag identification number for each PIT-tagged mussel. After all the mussels were marked, we placed all mussels back within the quadrat. Approximately 2-3 weeks later, we resurveyed the 5 m x 5 m sites again using a Biomark HPR Plus Reader to locate pit tagged mussels in the Neches River sites (*Neches 1*, *Neches 2*, and *Neches 3*) and the Angelina site (*Angelina 1*). After scanning the site we also excavated the 5 m x 5 m quadrat by hand with the use of the 1m² quadrats to locate any unmarked mussels and I recorded the tag number, tag color, and the number of unmarked target mussels.

Mark-Recapture Analysis

Mark-recapture models were created from the data for each location using Program MARK version 8.0 (White and Burnham 1999). Program MARK is used to provide parameter estimates from organisms that are marked and reencountered, whether alive or dead, and can provide population size estimates within closed populations (White and Burnham 1999). The POPAN model was chosen for this study because in addition to estimating capture probability, survival probability, and overall population size this model has the capability to estimate the population size at each encounter (Arnason and Schwarz 1995, Schwarz and Arnason 1996). I created models for each of the seven field sites and both real values and derived values were extracted from the top model(s) for each site. I recorded the real and derived estimates for ϕ_1 (survival probability between the initial visit and the 2nd visit), p_2 (probability of capture at 2nd visit), N_1 (initial population estimate), N_2 (population at 2nd visit), and Gross N (overall population estimate during course of study) for the top model(s) at each field site. These values are reflective of the population with the 5 m x 5 m mark-recapture areas and not the entire field site. In addition, I also calculated the recapture frequencies for each site visit using data collected during the mark-recapture study to determine any significant difference between visits and rivers.

0.25 m² Quadrat Survey Analysis and Size Classes

The data we collected from the initial 0.25 m² quadrat surveys was used to calculate mean m² densities and the densities of the mark-recapture sites by using the total number of individuals divided by the total mark-recapture area (25 m²). As two of the sites on the Sabine River had quadrat data collected in 2015 instead of 2014, I

compared the densities from the quadrat data to the densities of the mark-recapture site obtained during the first visit in 2015 (2nd mark-recapture visit). Initial quadrat survey data from all other sites were obtained in 2014 and were compared to mark-recapture site densities during the first visit in 2014. In addition to densities, I calculated the mean lengths of each species from both the 0.25 m² quadrat surveys and the mark-recapture sites and compared them to each other. T-tests were used to compare data collected by the different survey methods. The size classes were created by using the lengths of mussels sampled from both the mark-recapture sites and from the 0.25 m² quadrat surveys. Survey data from 2014 was pooled with mark-recapture lengths from 2014 and survey data from 2015 was pooled with mark-recapture lengths from 2015. Histograms were created with R Studio, with packages dplyr and ggplot2 (R Core Team 2013, Wickham and Francois 2013, Wickham 2009).

Results

POPAN Estimates

Model selection in MARK resulted in a total of 10 top models as three sites had two top models of equal values (Table 3.2). All ranked models created for each site are included in Appendix A. The largest population sizes were estimated for *F. askewi* populations (Table 3.2) Largest Gross N values were also estimated for *F. askewi* populations (Table 3.2) An increase in estimated population size was seen in most populations between 2014 and 2015 except at sites *Sabine 2* and *Neches 3* (Table 3.2). Though the largest population estimate from the models was for *F. askewi*, there were no significant difference in estimated population sizes between the Neches River and the

Sabine River within the 25 m² mark-recapture area. This can be seen both during 2014 and 2015 (Table 3.2). Though there was no significant difference in population sizes, the largest population size estimated for *F. askewi* was higher than the largest estimated population size for *P. riddellii*. The largest population estimate for a *F. askewi* population was at site *Sabine 1* with a population estimate of 302 ± 26.72 in 2015 while the largest population estimate for a *P. riddellii* population was at site *Neches 1* with a population estimate of 101 ± 4.99 in 2015 within the 25 m² mark-recapture area (Table 3.2). Both sites *Neches 3* and *Sabine 2* saw a decrease in estimated population sizes between 2014 and 2015 while all other sites saw an increase in estimated population size during this time within the 25 m² mark-recapture area (Table 3.2). Though site *Neches 3* saw a small decrease in estimated population size in 2015, site *Neches 3* had a larger gross N value than site *Neches 2*. A value of 91 ± 2.33 was calculated for site *Neches 3* and 71 ± 2.89 for site *Neches 2* (Table 3.2). This larger gross N value indicates that model estimated that there were more individual mussels present in the 25 m² mark-recapture area over the course of the study, which can also be seen when including the number of individuals found during the final visit (Table 3.3).

Table 3.2. Real and derived estimates with standard error values from top POPAN models for all seven mark-recapture sites. Neches 3, Sabine 2, and Angelina 1 had two models top models with equal real and derived estimates. These values are reflective of the 25 m² mark-recapture site and not the entire field site.

Site	Model*	Survival (ϕ_1)	Capture (p_2)	N_1 at 2014	N_2 at 2015	Gross N
<i>Pleurobema riddellii</i>						
Neches 1	$\phi(t)pent(t)p(t)$	0.92 ± 0.03	0.85 ± 0.06	64 ± 5.78	101 ± 4.99	150 ± 3.75
Neches 2	$\phi(t)pent(t)p(t)$	0.98 ± 0.01	1 ± 0	29 ± 3.96	60 ± 3.12	71 ± 2.89
Neches 3	$\phi(.)pent(t)p(t)$	0.98 ± 0.01	1 ± 0	58 ± 4.12	54 ± 4.21	91 ± 2.33
	$\phi(.)pent(t)p(.)$					
<i>Fusconaia askewi</i>						
Sabine 1	$\phi(.)pent(t)p(.)$	0.92 ± 0.01	0.44 ± 0.04	280 ± 34.56	302 ± 26.72	909 ± 97.03
Sabine 2	$\phi(t)pent(t)p(t)$	0.02 ± 0.07	1 ± 0	90 ± 6.12	64 ± 6.12	381 ± 63.95
	$\phi(t)pent(t)p(.)$					
Sabine 3	$\phi(.)pent(t)p(.)$	0.97 ± 0.01	0.44 ± 0.05	116 ± 20.19	169 ± 18.61	293 ± 32.75
<i>Fusconaia</i>						
Angelina 1	$\phi(t)pent(t)p(.)$	0.77 ± 0.04	1 ± 0	42 ± 4.3	95 ± 4.71	181 ± 13.49
	$\phi(.)pent(t)p(t)$					

* ϕ , survival probability; p , capture probability; $pent$, recapture probability; N , population estimates; $(.)$, constancy; (t) , temporal variation

Mean density values for *F. askewi* and *P. Riddellii*

Fusconaia askewi was found in the Neches, Sabine, and Angelina Rivers during 0.25 m² surveys (Table 3.3). Although *F. askewi* was found in all three rivers, width and height measurements for sampled *F. askewi* were only gathered at sites on the Sabine River. With the 0.25 m² surveys we did not detect *F. askewi*, live or dead, at site Sabine 2 though this site was still used as a mark-recapture site (Table 3.2, Table 3.3). Site Sabine 1 had the highest mean live m² density for *F. askewi* at 5.63 ± 5.90 and also had the highest mean m² density for dead *F. askewi* at 9.04 ± 10.26 . Site Sabine 3 during the summer of 2015 had a mean m² density for live *F. askewi* at 0.89 ± 2.56 (Table 3.3). No dead *F. askewi* were found at site Sabine 3 during quadrature surveys (Table 3.3). *P. riddellii* was found only in the Neches River during 0.25 m² surveys (Table 3.3). The

0.25 m² surveys did not detect *P.riddellii*, live or dead, at site *Neches 3* though this site was still used as a mark-recapture site (Table 3.2, Table 3.3). Site *Neches 2* had the highest mean m² density for live *Pleurobema riddellii* at 2.82 ± 7.34 and site *Neches 1* had the lowest mean m² density for live *P. riddellii* at 0.22 ± 0.51 (Table 3.3). Neither site had any dead individuals during the 0.25 m² surveys (Table 3.3).

Table 3.3. Mean m² densities for *Fusconaia askewi* and *Pleurobema riddellii* from surveys between 2014-2015.

<i>Site</i>	<i>Fusconaia askewi</i>		<i>Pleurobema riddellii</i>	
	<i>Mean m² density Live</i>	<i>Mean m² density Dead</i>	<i>Mean m² density Live</i>	<i>Mean m² density Dead</i>
<i>Neches 1</i>	0.54 ± 1.45	0.15 ± 0.70	0.222 ± 0.51	0 ± 0
<i>Neches 2</i>	4.37 ± 7.58	0.07 ± 0.54	2.815 ± 7.34	0 ± 0
<i>Neches 3</i>	0.22 ± 1.21	0.22 ± 0.93	--	--
<i>Sabine 1</i>	5.63 ± 5.90	9.04 ± 10.26	--	--
<i>Sabine 2*</i>	--	--	--	--
<i>Sabine 3*</i>	0.89 ± 2.56	0 ± 0	--	--
<i>Angelina 1~</i>	2.22 ± 9.35	0 ± 0	--	--

*Indicates sites surveyed during summer 2015. All other sites were surveyed during summer 2014.
 ~All individuals surveyed were identified as *Fusconaia askewi*

Comparison of size distributions over a 1-year period

I used lengths collected from both mark-recapture sites and from 0.25 m² quadrat surveys were used to create size classes for the mussels from all seven study sites, separated by year (Appendix B). Lengths obtained from the 0.25 m² quadrat surveys were pooled within the year they were collected. Overall, sites on the Sabine River exhibited a larger number of size classes compared to the Neches River. Site *Neches 1* saw a shift in lengths, with more individuals in 2015 being measured between 30-50 mm as opposed to 2014, which saw a larger number of individuals in the 20-40 mm range (Figure 3.2). Site

Neches 2 both years had the 40-50 mm size class as the most common group (Figure 3.3). Between 2014 and 2015 at site *Neches 2* also saw a decrease in the smallest size class and an increase in the largest size class (Figure 3.3). Site *Neches 3* had a size class of 0-10 mm that did not reappear in 2015 (Figure 3.4). Size classes at site *Sabine 1* skewed-left with more individuals being recorded in larger size classes in 2015 (Figure 3.5). *Sabine 2* saw a large shift in size distribution becoming left-skewed; smaller individuals were measured in 2014 and individuals measured in 2015 were larger (Figure 3.6). Though the larger sizes classes at *Sabine 3* did see an increase in the number of individuals measured in 2015, smaller individuals between 10-40 mm were found in 2015 and not seen in 2014 (Figure 3.7). All sizes at site *Angelina 1* saw an increase in number of individuals between 2014 and 2015, with a new 10-20 mm size class added in 2015 (Figure 3.8).

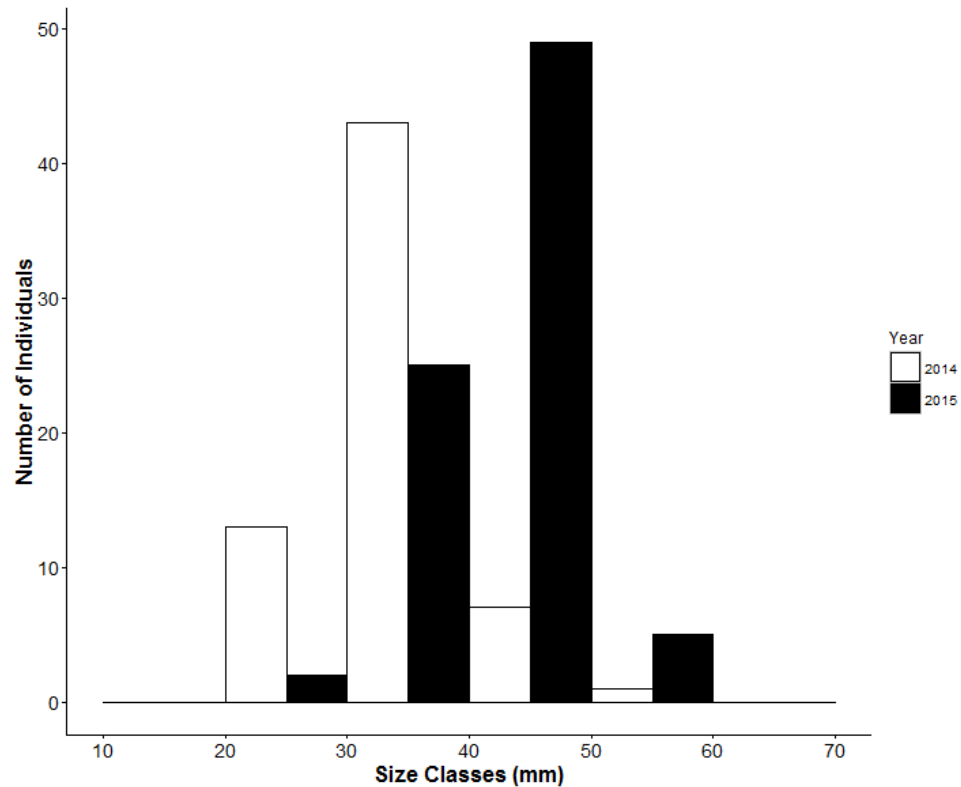


Figure 3.1. Sizes classes for *Pleurobema riddellii* at site *Neches 1*, by year.

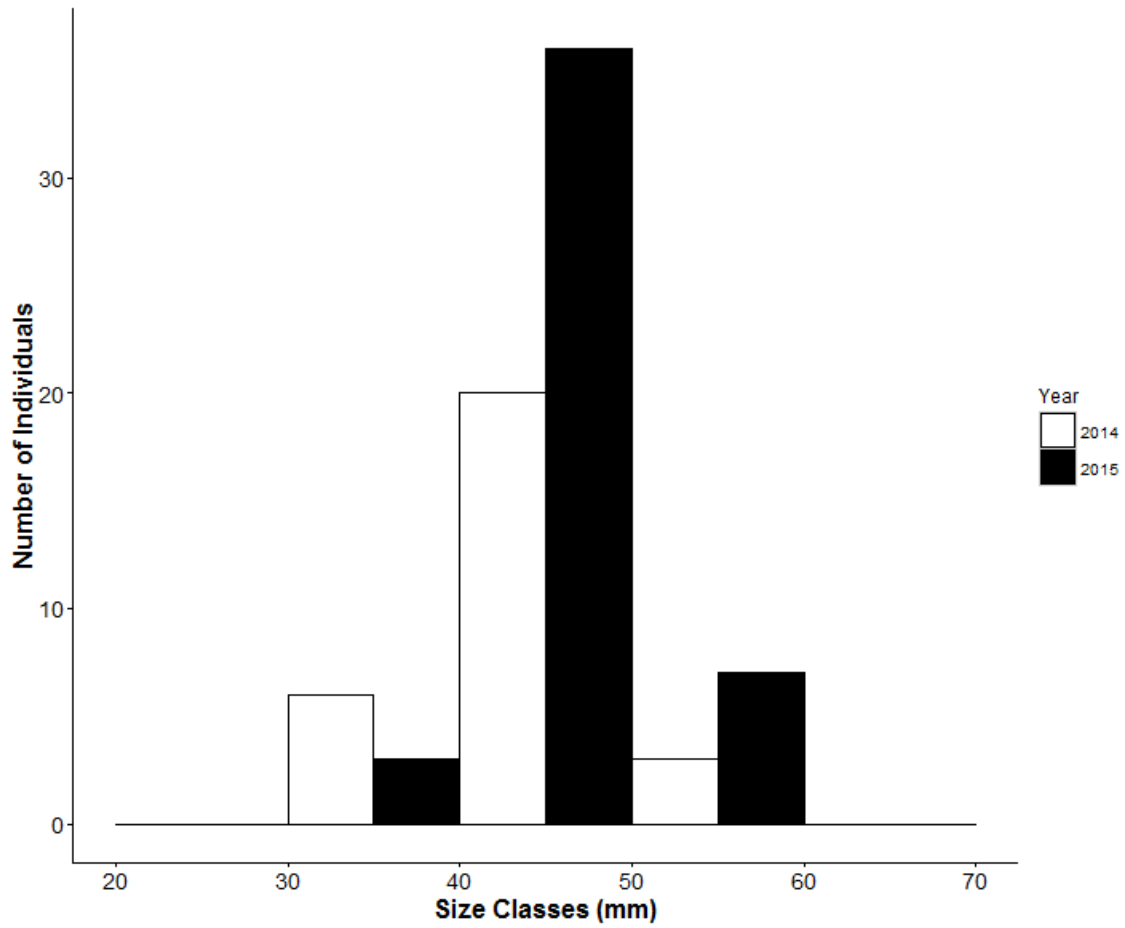


Figure 3.2. Sizes classes for *Pleurobema riddellii* at site *Neches 2*, by year.

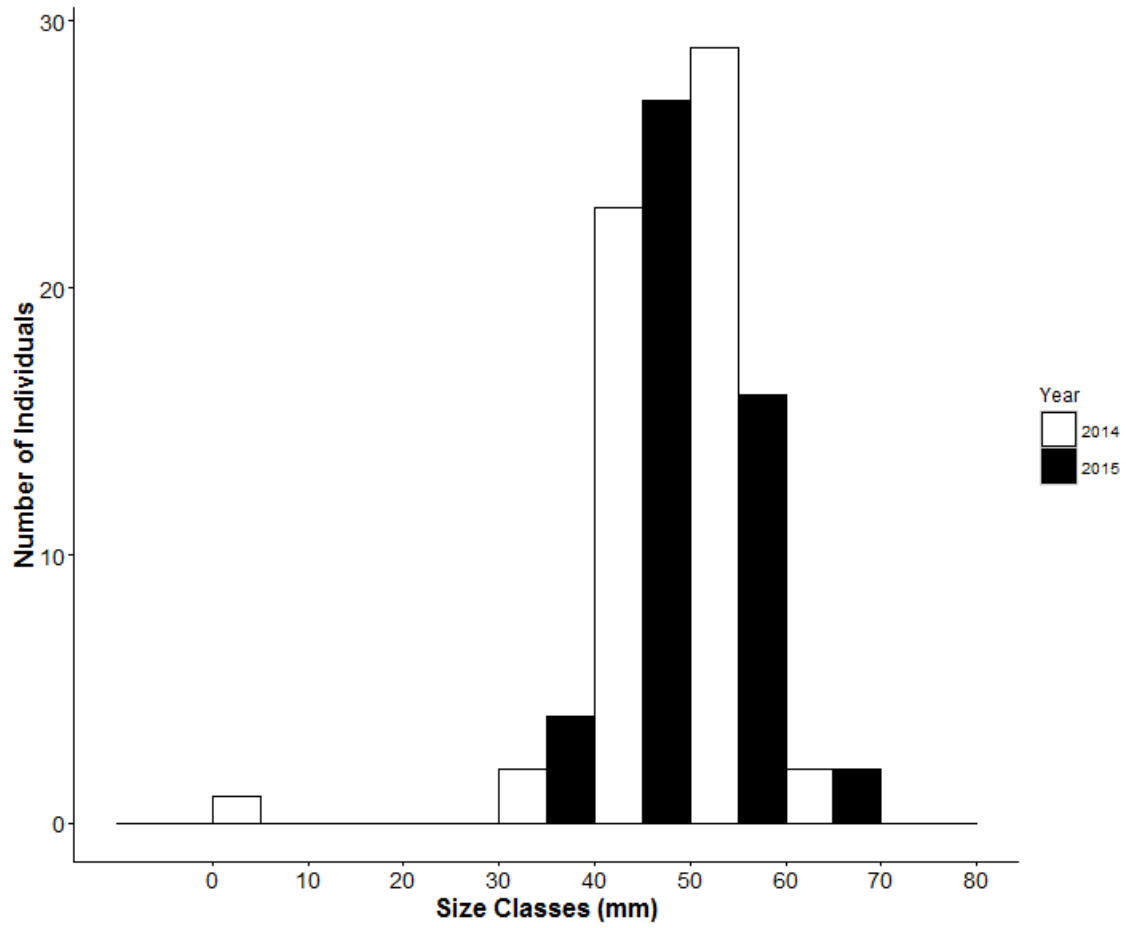


Figure 3.3. Sizes classes for *Pleurobema riddellii* at site *Neches 3*, by year.

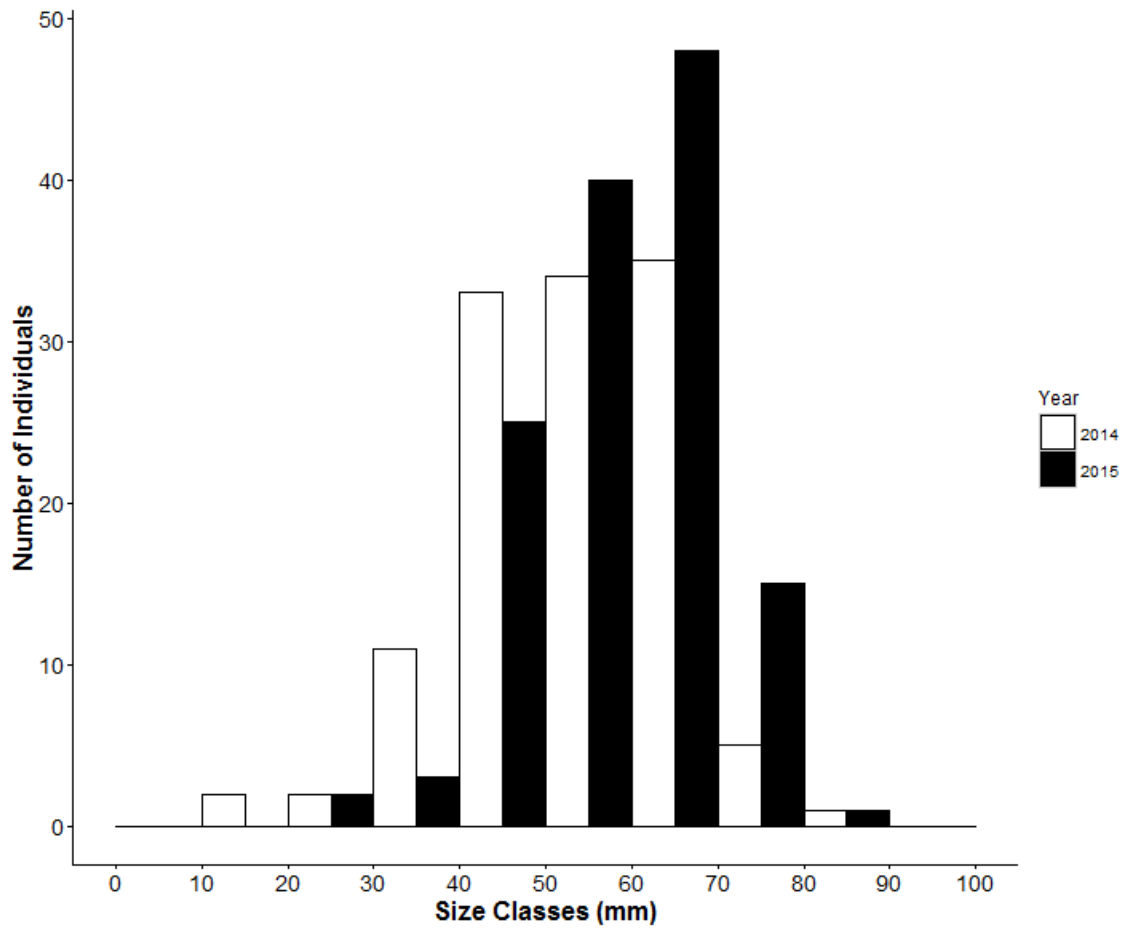


Figure 3.4. Sizes classes for *Fusconaia askewi* at site *Sabine I*, by year.

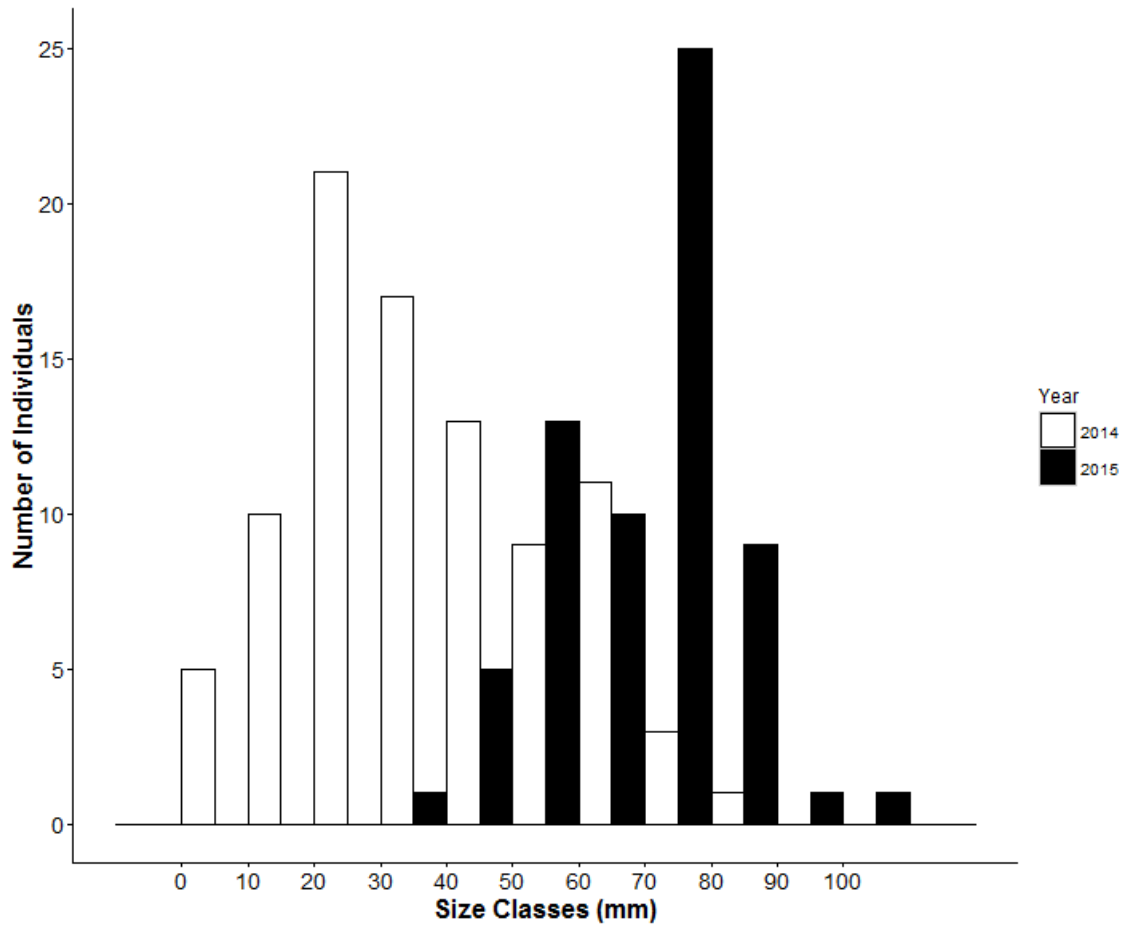


Figure 3.5. Sizes classes for *Fusconaia askewi* at site *Sabine 2*, by year.

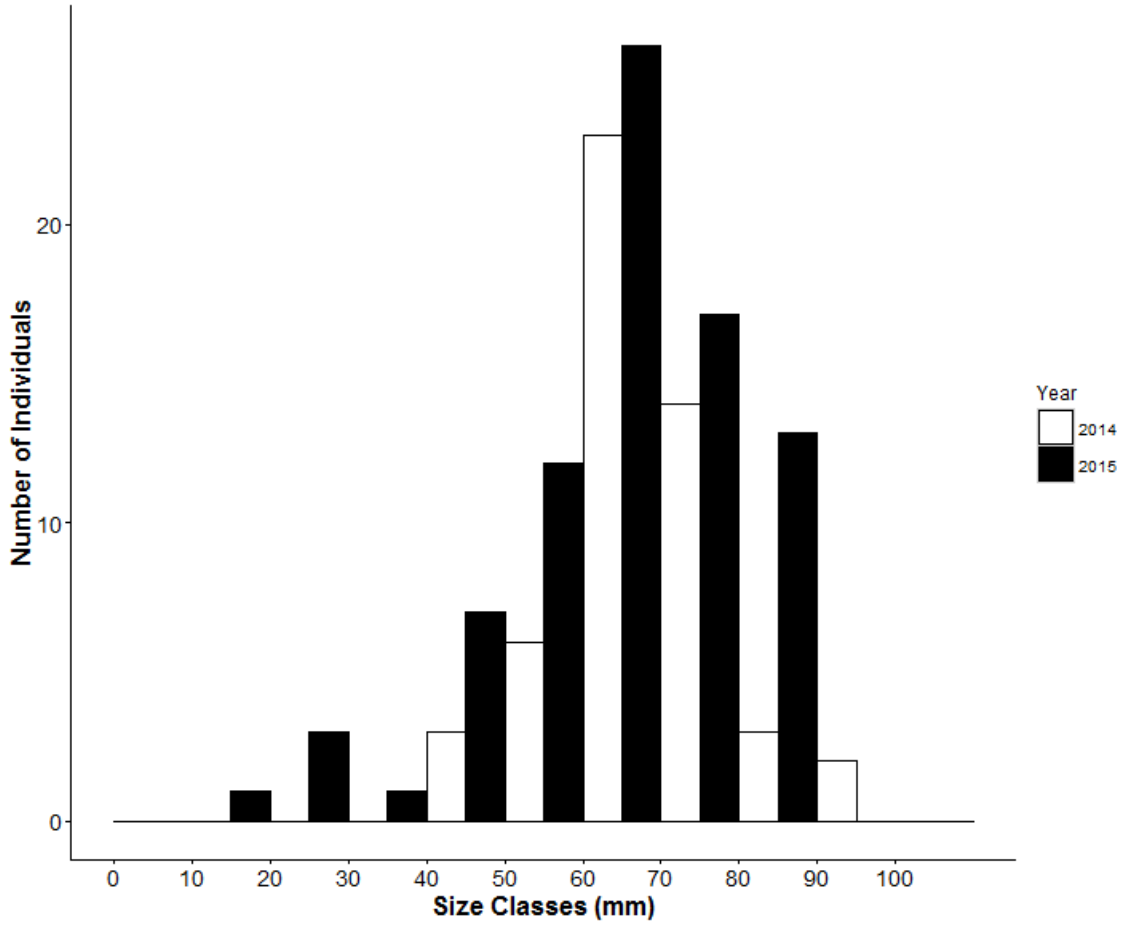


Figure 3.6. Sizes classes for *Fusconaia askewi* at site *Sabine 3*, by year.

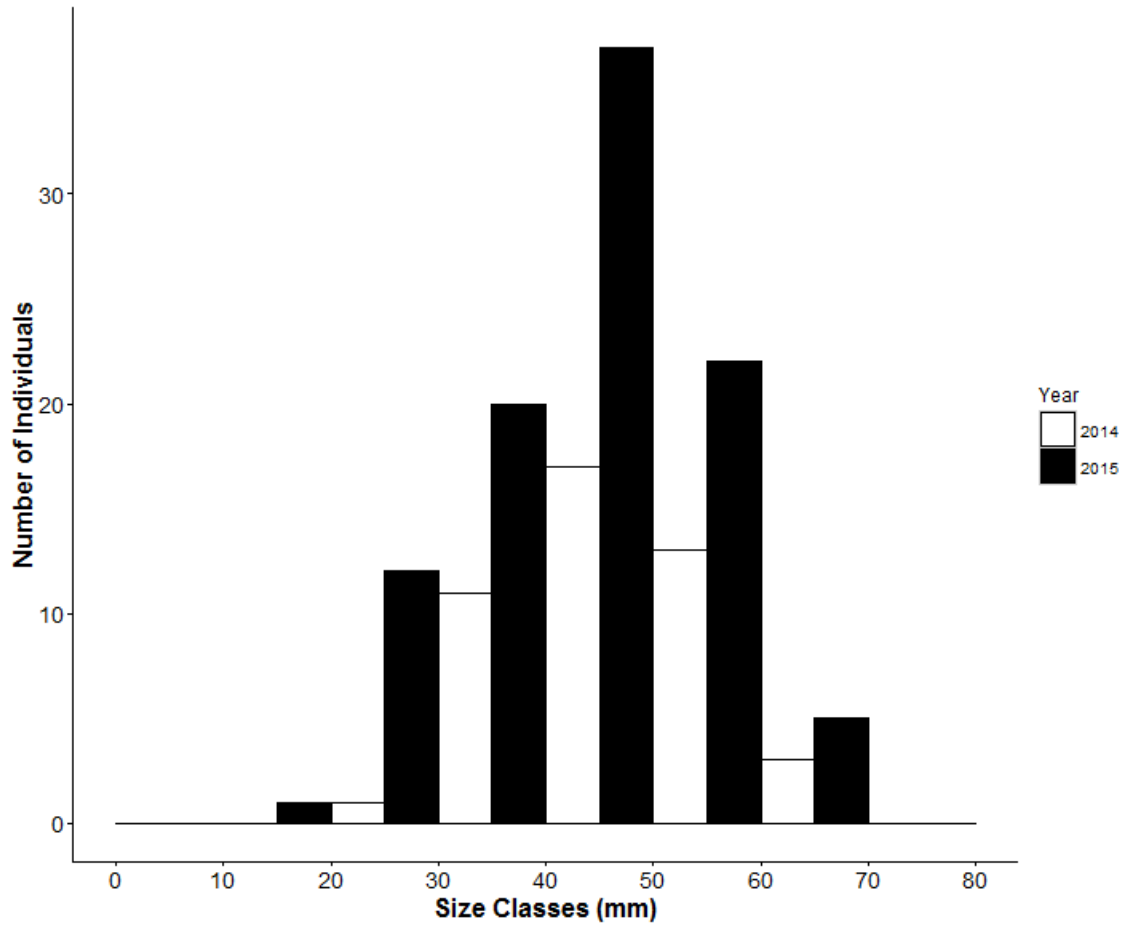


Figure 3.7. Sizes classes for *Fusconaia* at site *Angelina 1*, by year.

Comparison of recapture rates

Values obtained from the mark-recapture study were also used to calculate the percentages of individuals recaptured between the three site visits. All sites had a higher recapture rate between the 2-week periods in 2015 versus the one-year period between 2014-2015 (Table 3.4). All sites that had PIT tags added to the mussels during the 2nd site visit had significantly higher recapture values between the 2nd and 3rd site visits (Table 3.4; p-value < 0.05, t-value = 12.99, df = 4). All sites on the Neches River had a significantly larger recapture rate between the one-year period (1st and 2nd site visit) than the sites on the Sabine River (Table 3.4; p-value < 0.05, t-value = 9.33, df = 4). Additionally, all sites on the Neches River had a marginally significant increase in recapture rates when the PIT tags were added compared to the sites on the Sabine River where PIT tags were not used (Table 3.4; p-value < 0.06, t-value = 3.99, df = 4). Approximately 9% of mussels at site *Angelina 1* were identified as *F. lananensis* during 2014 based on external morphology while all other specimens captured were classified as *Fusconaia* (Table 3.4).

Table 3.4. Recapture Rates and Total Captured for all seven mark-recapture sites within the 25 m² sites.

<i>Site</i>	<i>1st Visit, 2014</i>	<i>Recapture Rate between Visit 1 and 2</i>	<i>2nd Visit, 2015 Total</i>	<i>Recapture Rate between Visit 2 and 3</i>	<i>3rd Visit, 2015 Total</i>
<i>Neches 1*</i>	63	0.49	77	1	100
<i>Neches 2*</i>	28	0.61	45	1	50
<i>Neches 3*</i>	58	0.55	56	1	61
<i>Sabine 1</i>	122	0.07	125	0.49	126
<i>Sabine 2</i>	89	0.01	64	0.02	2
<i>Sabine 3</i>	50	0.14	72	0.53	65
<i>Angelina 1*</i>	40	0.33	81	0.94	76

*Indicates sites where all individuals in 2015 were marked with PIT Tags in addition to bee tags

Quadrat densities versus mark recapture densities

Mean m^2 densities were also calculated for all mark-recapture sites for comparison to the values calculated from the $0.25 m^2$ surveys (Table 3.6). Estimates of density for the mark-recapture portion of the study were calculated by dividing the total number of individuals found during the visit divided by $25 m$ (Table 3.5). There was no significant difference in m^2 density between the use of $0.25 m^2$ quadrats or mark-recapture sites (Table 3.5). Higher densities were found using $0.25 m^2$ quadrats for sites Neches 2, Sabine 1, and Angelina 1 while other sites had higher values calculated from mark-recapture data (Table 3.5). In addition to mean densities, mean lengths of individuals sampled during both $0.25 m^2$ quadrat surveys and mark-recapture sites were compared (Table 3.6). There was no significant difference in mean lengths between mussels in the $0.25 m^2$ quadrat surveys or mark-recapture sites (Table 3.6). Greater mean lengths were found at sites Neches 1 and Sabine 1 through the use of $0.25 m^2$ quadrats while other sites had higher values calculated through the mark-recapture study (Table 3.6).

Table 3.5. Comparison of mean m² densities between quadrat surveys and mark-recapture values. Mean densities for mark-recapture sites were obtained by dividing the total captured by 25 m.

<i>Site</i>	<i>Mean m² density</i>	
	<i>m² Quadrats</i>	<i>Mark-Recapture</i>
<i>Pleurobema riddellii</i>		
<i>Neches 1</i>	0.22±0.51	2.52
<i>Neches 2</i>	2.82±7.34	1.12
<i>Neches 3</i>	0	2.32
	Mean for <i>Pleurobema riddellii</i>	
	1.01	1.99
<i>Fusconaia askewi</i>		
<i>Sabine 1</i>	5.63±5.90	4.88
<i>Sabine 2*</i>	0±0	2.52*
<i>Sabine 3*</i>	0.89±2.56	3.16*
	Mean for <i>Fusconaia askewi</i>	
	2.17	3.52
<i>Fusconaia</i>		
<i>Angelina 1</i>	2.22±9.35	1.76

*Indicates sites where 0.25² surveys were conducted during summer 2015. Mark-recapture density values for these sites reflect values obtained during 2015. All other values reflect the number captured during initial visits in 2014.

Table 3.6. Comparison of lengths (mm) of individuals measured during quadrat surveys and mark-recapture. Includes both mean and range for each method.

<i>Site</i>	<i>Lengths (mm)</i>			
	<i>m² Quadrats</i>		<i>Mark-Recapture</i>	
	Mean	Range	Mean	Range
<i>Pleurobema riddellii</i>				
<i>Neches 1</i>	42.30±3.68	36.55-47.00	34.3±5.86	21.6-53.3
<i>Neches 2</i>	38.63±6.18	28.2-51.1	44.7±4.12	36.2-53.3
<i>Neches 3</i>			48.5±8.05	38.7-63.3
<i>Fusconaia askewi</i>				
<i>Sabine 1</i>	54.23±15.88	12-85	53±11.69	16.1-85.8
<i>Sabine 2</i>	46.85±13.91	14.8-79.4	68.7±12.75	42.5-104.3
<i>Sabine 3</i>			64.1±15.29	16.5-88.1
<i>Fusconaia</i>				
<i>Angelina</i>	43.57±10.51	31.7-64.4	45.7±8.67	29.9-69.9

*Indicates sites where 0.25² surveys were conducted during summer 2015. Mark-recapture length values for these sites reflect values obtained during 2015. All other values reflect the lengths recorded during initial visits in 2014.

~Indicates sites where no individuals were captured during 0.25 m² quadrat surveys.

Discussion

Population dynamics at high density sites

Size classes suggest that *P. riddellii* may reach an adult size between 30-40mm, as this size class had the largest number of individuals (Figure 3.1, Figure 3.2, Figure 3.3). Therefore I suggest *P. riddellii* below a length of 30 mm may be considered juveniles although this should be analyzed with assessment of gametes in the gonads. The population of *P. riddellii* at site *Neches 3* had a greater range of size classes than other sites on the Neches River (Figure 3.3). The *P. riddellii* populations at site *Neches 1* and *Neches 3* had individuals less than 30 mm in length, indicating some recruitment at these sites (Figure 3.1, Figure 3.3). The lack of size classes smaller than 30 mm for *P. riddellii* populations could also be the result of juveniles burrowing, which would decrease our detection of juveniles (Figure 3.1, Figure 3.2, Figure 3.3, Vaughn and Hakenkamp 2001).

For *F. askewi*, individuals appear to reach an adult size after reaching approximately 40 mm (Figure 3.4, Figure 3.5, Figure 3.5, Figure 3.7). Therefore I suggest *F. askewi* below 40 mm may be considered juveniles although this should be analyzed with assessment of gametes in the gonads. Size distributions for the *F. askewi* in the Sabine River showed a wide diversity in size classes even though mark-recapture studies have been known to be biased against smaller individuals, *Fusconaia askewi* populations at three field sites had shifts in the second year to larger size classes which could be due, in part, to the growth of individuals over the one year period and the immigration of new adults into the site (Figure 3.4, Figure 3.5). Overall, my *F. askewi* populations had lower recapture rates between 2014-2015 than my *P. riddellii* populations (Table 3.2). The *F.*

askewi population at site *Sabine 1* had small size classes that were detected again in 2015; it is possible these individuals grew into the next size classes, were dislocated or killed during the flooding, or burrowed further into the sediment (Figure 3.4). The *F. askewi* population at site *Sabine 3* gained additional sizes classes in 2015; this addition of smaller size classes indicates recruitment may have occurred in the population at *Sabine 3* (Figure 3.6). It is also possible that the new juveniles were moved into the study site during flooding. Overall sites *Sabine 1* and *Sabine 2* had almost normal distributions that were skewed slightly left, which would be expected of a population with a uniform distribution of a longer lived species (Miller and Payne 1993, Haag and Warren 2007, Haag and Warren 2010, Haag 2012).

The largest population sizes within the mark-recapture areas were found in the Sabine and Angelina Rivers where *F. askewi* populations were tracked (Table 3.2). In addition, the *F. askewi* population at *Sabine 1* also had a large gross N value of 909 ± 97.03 (Table 3.2). This difference in population estimates for these species agreed with previous research where *F. askewi* was found in much higher densities than *P. riddellii* (Burlakova et al. 2010, Ford et al. 2012, Ford et al. 2014). Unlike other surveys, no difference was found between the mean density of *F. askewi* in the Neches River or the Sabine River. *F. askewi* was typically found in higher densities in the Sabine River (Ford et al. 2012). More apparent juveniles for species *F. askewi* were detected than for *P. riddellii* during the study. Indeed the lack of small size classes suggest recruitment may not be high with these *P. riddellii* populations but is occurring with the *F. askewi* populations, particularly site *Sabine 2* (Figure 3.5, Nalepa and Gauvin 1988, Haag 2012).

Average m² at high density sites: comparisons to other studies

Mean m² densities calculated for *P. riddellii* were 1.99 per m² and mean densities for *F. askewi* were 3.52/m² (Table 3.5). Surveys for Texas threatened species *Popenaias popeii*, the Texas Hornshell, calculated densities were between 0-0.186 individuals per m² (Karatayev et al. 2015). Species density for a locally rare species *Quadrula pustulosa*, the Pimpleback, in Wisconsin had estimates of approximately 0.25 individuals per m² (Sethi et al. 2004). In comparison, common mussel species had values of up to 1.25 individuals per m² in the same area (Sethi et al. 2004). When surveys were conducted in locations in French Creek, Ohio, *Epioblasma torulosa rangiana*, the Northern Rifle Shell, had m² densities ranging from 0.01 – 6.67 m² (Crabtree and Smith 2009). When comparing to those found in other studies, both *P. riddellii* and *F. askewi* had higher densities than expected for a rare mussel species. The density values obtained for *P. riddellii* and *F. askewi* are more comparable to densities one might expect for common mussel species, not rare mussel species. These higher values for *P. riddellii* and *F. askewi* should be expected as my study were conducted at sites with apparently high suitable habitat that historically had higher densities of these species. Quadrat surveys for these species at random site locations would likely produce densities similar to those found in other studies involving rare species.

Effect of flooding on recapture rates in the Sabine River

Between the 1st and 2nd mark-recapture visits, 2014 and 2015, the mussels at Neches River sites had higher recapture rates than sites on the Sabine River (Table 3). One significant event that occurred in the Sabine River between 2014 and 2015 was heavy flooding. Both the Sabine River and the Neches River experienced higher water

levels during the winter of 2014-2015 as opposed to the winter of 2013-2014 (Figure 3.9, Figure 3.10, Figure 3.11, Figure 3.12). Data from gauges on both rivers near the mark-recapture sites show that the sites on the Sabine River measured more flooding during the 1st and 2nd visits than sites on the Neches River (Figure 3.9, Figure 3.11). Flooding has been found to have negative effects on mussel population, including killing a significant portion of the population (Strayer 1999, Hastie et al. 2001). Flooding may have impacted mussels in the Sabine River; however, the number of *F. askewi* found at these sites in 2015 were similar to number found before the flooding event. Potentially, the flooding dislocated the marked *F. askewi* out of the mark-recapture areas, resulting in the smaller recapture rates in the Sabine River sites.

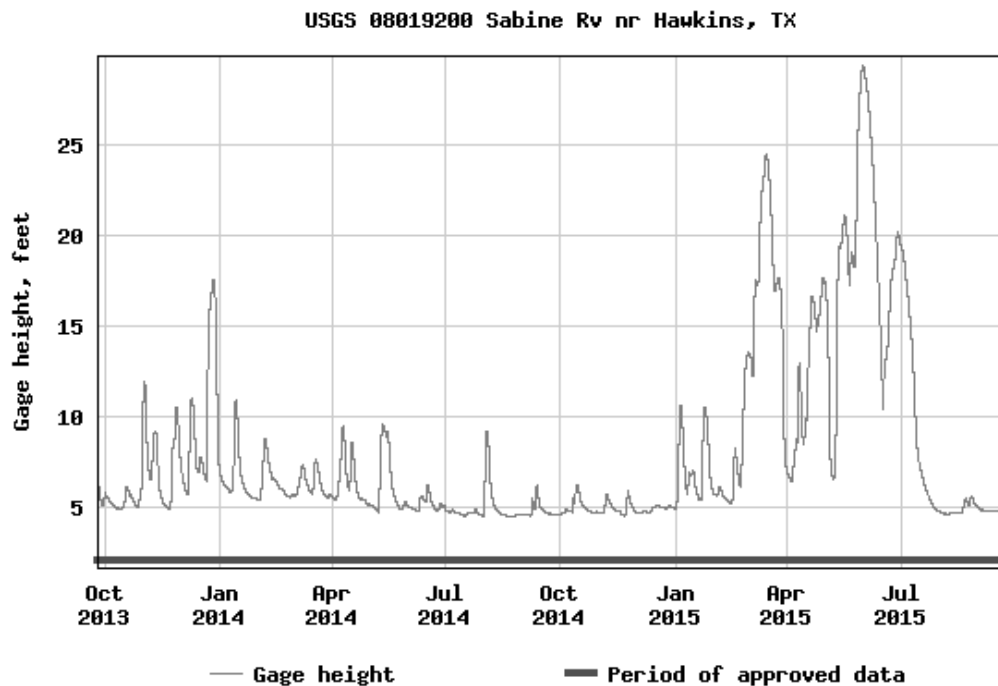


Figure 3.8. Gauge for the Sabine River in Hawkins, TX upstream from sites Sabine 1 and Sabine 2. Date ranges are from September 2013 through September 2015. Graph courtesy of the USGS.

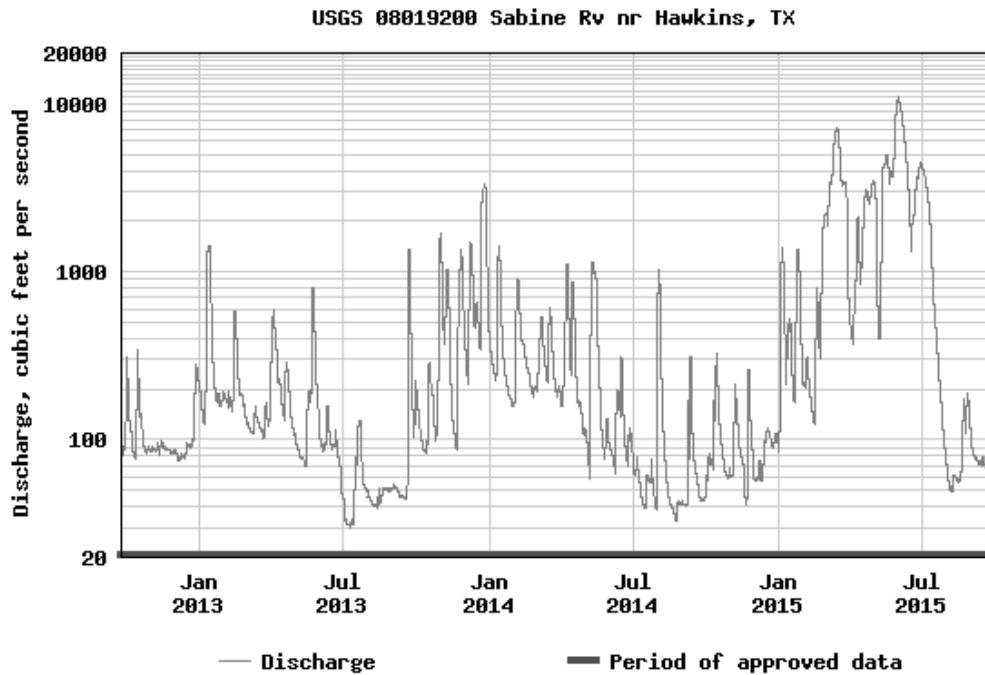


Figure 3.9. Discharge for the Sabine River in Hawkins, TX upstream from sites Sabine 1 and Sabine 2. Date ranges are from September 2013 through September 2015. Graph courtesy of the USGS.

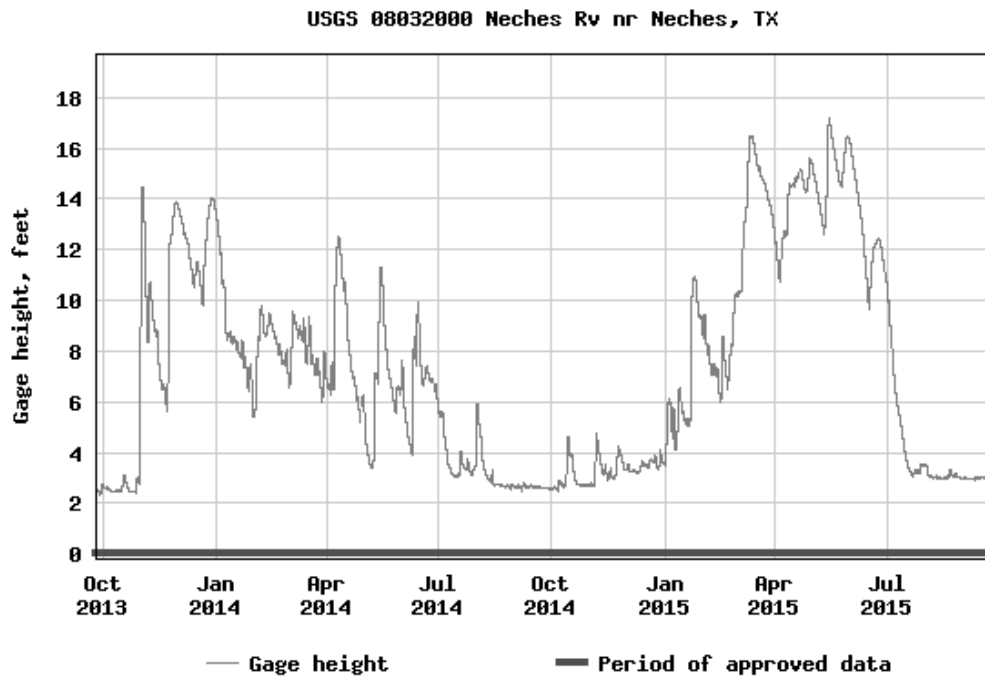


Figure 3.10. Gage heights for Neches River outside Neches, TX near site Neches 3. Date ranges are from September 2013 through September 2015. Graph courtesy of the USGS.

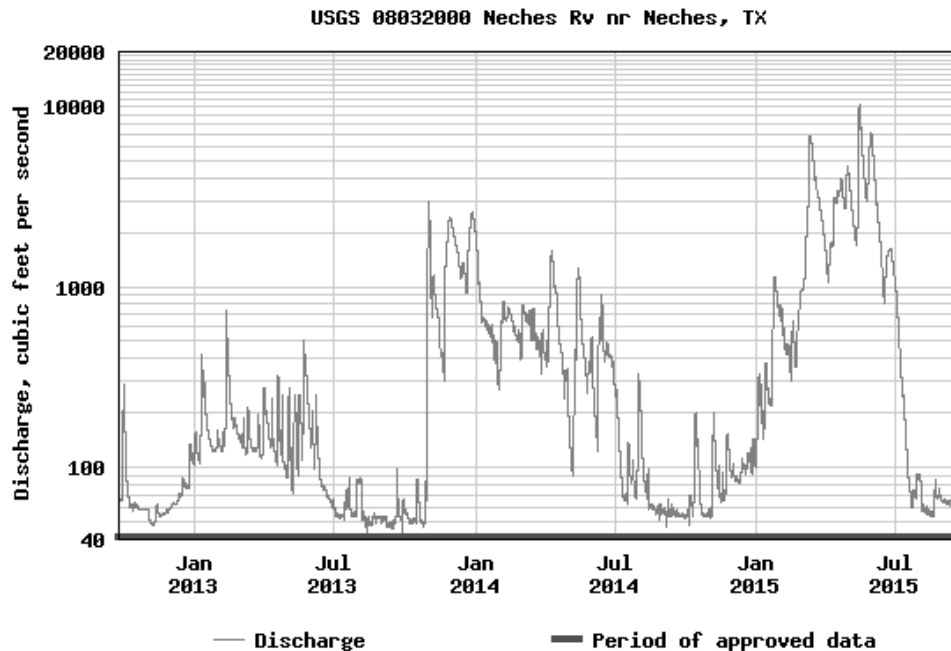


Figure 3.11 Discharge for Neches River outside Neches, TX near site Neches 3. Date ranges are from September 2013 through September 2015. Graph courtesy of the USGS.

Site Sabine 2

One site on the Sabine River had survival rates much lower than initially anticipated for *F. askewi*. Site *Sabine 2* has been previously used in other studies and has also been used in class field trips because of easy accessibility and the abundance of freshwater mussels, particularly *F. askewi* (Bakken 2013). However a low number of *F. askewi* were recaptured during each return visit and the top POPAN model estimated a low survival rate of 0.02 ± 0.02 , much lower than all other field sites (Table 3.2, Table 3.3). The top models for this *F. askewi* population also showed temporal variation with the survival probability (Table 3.2). In addition, size class distribution changed dramatically between 2014 and 2015, changing from a right-skewed distribution to a left skewed distribution, i.e. changing from a juvenile dominated distribution to an adult dominated distribution (Figure 3.5). This *F. askewi* population also had a right-skewed

during the year 2014; this type of distribution is typically not seen in other studies freshwater mussels (Bauer 1983, Hastie et al. 2000, Rogers et al. 2001, Haag 2012). This type of distribution is often seen with short-lived species; however, as *F. askewi* is believed to be a longer-lived species this type of distribution could indicate high recruitment in this population (Crabtree and Smith 2009, Haag and Warren 2010, Jones and Neves 2011, Haag 2012). Distribution changed dramatically in 2015 with the smallest size classes disappearing, likely related to the large amount of flooding between 2014 and 2015 (Figure 3.8, Figure 3.9).

Like the other sites on the Sabine River, *Sabine 2* experienced a large amount of flooding between the one-year mark-recapture time period between 2014-2015 (Table 3.2, Figure 3.8, Figure 3.9). The *F. askewi* populations at sites *Sabine 1* and *Sabine 3* did not have a dramatic drop in survival probability (Table 3.2, Figure 3.8, Figure 3.9). Additionally *F. askewi* at site *Sabine 2* also had a very low recapture rate between the 2nd and 3rd site visit, much lower than other recapture rates during this time period for other *F. askewi* populations (Table 3.4). It is possible that the effects of the flooding were stronger at this particular site than the other *F. askewi* populations. Yet Site *Sabine 2* is about 0.7 km upstream from site *Sabine 3* and *Sabine 3* had a higher recapture rate than site *Sabine 2* between the 2nd and 3rd visit (Table 3.1, Table 3.4). Site *Sabine 2* is located a short distance downstream from a bridge, about 200 m, which may have had an effect on the *F. askewi* population during the study. Mussel abundance and bank stability have both been observed declining immediately downstream of bridges (Levine et al. 2003).

Effect of PIT Tags on recapture rates

The *P. riddellii* populations and the *F. askewi* population on the Angelina River all had a top model where p , or the capture probability, was time dependent while all sites on the Sabine River had a model where capture probability was constant (Table 3.2). This was likely because of the addition of the PIT tags at the Neches River and the Angelina River sites as these sites showed a marginally significant increase in recapture rates between the 2nd and 3rd visit (Table 3). All *F. askewi* populations on the Sabine River, where PIT tags were not used, did not show this amount of increase between the 2nd and 3rd visit (Table 3.4). This indicates that the addition of the PIT tags likely influenced the recapture probability on the Neches River sites and the Angelina Site. It should also be noted that during the period where only bee tags were used at all sites, between the 1st and 2nd visits, the Neches River already had significantly higher recapture rates than the Sabine River (Table 3.4). Again, the Sabine River had more flooding during this time period, which may have decreased recapture rates (Figure 3.8, Figure 3.9).

Comparison of quadrat densities and mark recapture densities

Mean m^2 densities and mean lengths were calculated for both the 0.25 m^2 quadrat surveys and the mark-recapture surveys (Table 3.5, Table 3.6). No difference was found between the m^2 densities or mean length of individuals found through the 0.25 m^2 quadrat surveys and the mark-recapture project, for both *P. riddellii* populations and the *F. askewi* (Table 3.5). Quadrat sampling has been known to underestimate the abundance of rare freshwater mussel species (Vaughn et al. 1997). In this study 0.25 m^2 quadrat surveys estimated similar densities to the mark-recapture sites. No difference between mean lengths were found between 0.25 m^2 quadrat surveys and the mark-recapture

surveys (Table 3.5). This may relate to the fact that these sites were high density sites for the species relative to overall distributions in the rivers. Though the mark-recapture sites in this study did track smaller individuals, mark-recapture studies involving freshwater mussels can be biased towards larger individuals in the population (Miller and Payne 1988, Haag and Warren 2007, Haag 2009). The handling of mussels during these studies have also negatively impact growth rates of handled mussels, leading to overestimates in age and size in mark-recapture studies (Haag and Commens-Carson 2008, Haag 2009). In contrast, quadrat sampling has been known to be an effective way to sample for smaller individuals such as juveniles, especially when compared to timed surveys (Amyot and Downing 1991, Vaughn et al. 1997). In this study the number of individuals collected during the 0.25 m² quadrat surveys was smaller than the number of individuals collected from the mark-recapture sites. It is possible that employing a larger number of 0.25 m² quadrats would have increased the probability of finding smaller individuals within these locations. In addition timing of the surveys may have biased the size of individuals measured, as juveniles of some species have been known to descend into deeper substrate during mid- to late summer (Amyot and Downing 1991, Vaughn et al. 1997).

Population dynamics: Conclusions

Overall, my *P. riddellii* populations and the *F. askewi* populations showed differences in population responses over the course of this study. The *F. askewi* populations had strong evidence of recruitment and a wide diversity of size classes; in contrast, the *P. riddellii* populations had little diversity in size class and little evidence of recruitment. Although the *P. riddellii* populations did not have evidence of recruitment occurring, the *P. riddellii* populations were much more stable than the *F. askewi*

populations in the Sabine River over the course of this study. The flooding that occurred during this study had a stronger impact on the population structure of the *F. askewi* populations than the *P. riddellii* populations. Flooding dislocated the original *F. askewi* populations on the Sabine River; however, though the new individuals present in the population had strong signs of recruitment and high abundance.

Chapter 4: Conclusions and Future Research

Additional molecular genetic analysis supports the conclusion that *Fusconaia lananensis* and *F. askewi* should not be designated as separate species. Along with the strong morphological similarities I suggest that *F. lananensis* and *F. askewi* should be combined into one species. However, as the specimens in the study were only from two locations it would be useful to analyze other populations.

The models I created provided estimates of population size at three sites for each species in addition to survival, capture, and entry probabilities for species *P. riddellii* and *F. askewi*. Though the population estimates are from 25 m² areas with high densities, with additional data recapture probabilities from other locations the models could be used to estimate population sizes over larger stretches of river (Inoue et al. 2014) which is the critical information needed for protection of threatened mussels. The use of PIT tags greatly increased the recapture rates, though the cost for this tagging method may remain an issue for some time. Bee tags may be better suited for short-term studies as they are easily applied and last for at least a year.

As expected, population sizes within the 25 m² mark-recapture areas for *P. riddellii* were lower than population estimates for *F. askewi*. Though earlier studies have found locations with large numbers of *F. askewi*, *F. askewi* is endemic to the Sabine, Neches, and Angelina Rivers (Ford et al. 2012). In comparison to the mean m² densities of rare mussels sampled during studies, the mean densities for *P. riddellii* and *F. askewi*, had values comparable to more common species than rare species (Table 3.5, Sethi et al. 2004, Karatayev et al. 2015). An explanation of these higher densities for *P. riddellii* and *F. askewi* is that these studies were conducted at sites that historically had the highest densities of these species. Quadrat surveys for these species at random site locations would more likely produce densities similar to those found in other studies involving rare species. Though *P. riddellii* have been located in several sites in the Neches River, *P. riddellii* is extremely rare in other river systems (Ford et al. 2012, Ford et al. 2014).. In addition, the results of the size distributions suggest little evidence of recruitment for *P. riddellii*; however further steps should be taken to increase detection of juveniles for species *P. riddellii*.

As conservation efforts for freshwater mussels increase, continued analysis of established freshwater mussel populations will be crucial to determine population status. *Fusconaia askewi* appears to be stable in the Sabine River because of the abundance of both adults and juveniles; however, *F. askewi* is endemic to this region (Ford et al. 2012) and so may merit continued protection considering it is less abundant in other river systems. The *P. riddellii* at all three Neches sites were marked with PIT tags during the summer of 2015 so these populations can continue to be monitored to gather more information for this species. It should be note that because these sites were chosen based

on historical sites with high density, the measures obtained from this study cannot be extrapolated to the entire population of river system for each species. Additional surveys in random river locations and more detailed research on mesohabitats would be needed before population level information could be extrapolated for the entire species.

However, the sites set up in this study can continue to be monitored to gather information at established sites at areas with historically high densities of these species. If the sites for *P. riddellii* are monitored, the population size and the size class distributions can be measured from year to year to watch for changes in population size or distribution.

Detection of juvenile *P. riddellii* will be essential to provide evidence for recruitment at these sites; if *P. riddellii* is a species with burrowing juveniles, additional intensive sampling methods may be needed. As these sites are representative of high density sites for *P. riddellii*, revisiting these sites in the future may be useful for additional studies involving these species, including topics such as reproductive seasonality and other life history traits.

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Appendix A: POPAN Model Rankings

Table A.1 Ranking of POPAN models for site *Neches 1*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(t)pent(t)p(t)$	188.58	0	0.57	1	5
$\varphi(.)pent(t)p(t)$	190.26	1.68	0.25	0.43	5
$\varphi(t)pent(t)p(.)$	192.31	3.72	0.09	0.15	5
$\varphi(.)pent(t)p(.)$	193.07	4.49	0.06	0.1	5
$\varphi(t)pent(.)p(t)$	194.16	5.58	0.03	0.06	6
$\varphi(.)pent(.)p(.)$	90595.57	90406.99	0	0	3
$\varphi(t)pent(.)p(.)$	90599.64	90411.06	0	0	5
$\varphi(.)pent(.)p(t)$	-90599.64	0	0	0	0

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.2 Ranking of POPAN models for site *Neches 2*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(t)pent(t)p(t)$	53.45	0	0.71	1	3
$\varphi(.)pent(t)p(t)$	55.27	1.82	0.29	0.403	3
$\varphi(t)pent(t)p(.)$	41020.94	40967.49	0	0	1
$\varphi(.)pent(t)p(.)$	41022.76	40969.31	0	0	1
$\varphi(t)pent(.)p(.)$	41023.01	40969.56	0	0	2
$\varphi(.)pent(.)p(t)$	41024.83	40971.38	0	0	2
$\varphi(.)pent(.)p(.)$	41024.83	40971.37	0	0	2
$\varphi(t)pent(.)p(t)$	41029.41	40971.96	0	0	5

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.3 Ranking of POPAN models for site *Neches 3*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(.)pent(t)p(t)$	99.15	0	0.41933	1	2
$\varphi(.)pent(t)p(.)$	99.15	0	0.41933	1	2
$\varphi(t)pent(t)p(.)$	101.12	1.98	0.1561	0.3723	3
$\varphi(t)pent(t)p(t)$	107.91	8.77	0.00524	0.0125	4
$\varphi(.)pent(.)p(t)$	132.42	33.27	0	0	2
$\varphi(t)pent(.)p(t)$	82175.96	82076.82	0	0	3
$\varphi(.)pent(.)p(.)$	82176.06	82076.92	0	0	3
$\varphi(t)pent(.)p(.)$	82178.06	82078.92	0	0	4

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.4 Ranking of POPAN models for site *Sabine 1*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(.)pent(t)p(.)$	287.29	0	0.363	1	4
$\varphi(t)pent(t)p(t)$	288.17	0.89	0.233	0.6409	5
$\varphi(.)pent(t)p(t)$	288.17	0.89	0.233	0.6409	5
$\varphi(t)pent(t)p(.)$	288.79	1.51	0.171	0.4698	5
$\varphi(.)pent(.)p(t)$	172454.4	172167.13	0	0	3
$\varphi(t)pent(.)p(.)$	172515.1	172227.79	0	0	5
$\varphi(.)pent(.)p(.)$	172607.3	172320.06	0	0	3
$\varphi(t)pent(.)p(t)$	NONE	NONE	NONE	NONE	NONE

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.5 Ranking of POPAN models for site *Sabine 2*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(t)pent(t)p(t)$	37	0	0.49357	1	4
$\varphi(t)pent(t)p(.)$	37	0	0.49357	1	4
$\varphi(.)pent(t)p(t)$	44.3	7.29	0.01286	0.0261	5
$\varphi(.)pent(t)p(.)$	113.1	76.1	0	0	4
$\varphi(t)pent(.)p(t)$	113.31	76.31	0	0	3
$\varphi(.)pent(.)p(t)$	127288.69	127251.6	0	0	2
$\varphi(t)pent(.)p(.)$	127326.87	127289.8	0	0	3
$\varphi(.)pent(.)p(.)$	127415.54	127378.5	0	0	3

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.6 Ranking of POPAN models for site *Sabine 3*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(.)pent(t)p(.)$	219.27	0	0.32062	1	4
$\varphi(t)pent(t)p(t)$	219.85	0.58	0.24014	0.749	5
$\varphi(.)pent(t)p(t)$	219.85	0.58	0.24014	0.749	5
$\varphi(t)pent(t)p(.)$	220.22	0.95	0.19911	0.621	5
$\varphi(.)pent(.)p(t)$	70628.4	70409.17	0	0	5
$\varphi(t)pent(.)p(.)$	72191.5	71972.3	0	0	4
$\varphi(.)pent(.)p(.)$	72212.5	71993.27	0	0	3
$\varphi(t)pent(.)p(t)$	NONE	NONE	NONE	NONE	NONE

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Table A.7 Ranking of POPAN models for site *Angelina 1*.

<i>Model*</i>	<i>AICc</i>	<i>Delta AICc</i>	<i>AICc Weight</i>	<i>Model Likelihood</i>	<i>Parameters</i>
$\varphi(t)pent(t)p(.)$	166.08	0	0.42486	1	3
$\varphi(.)pent(t)p(t)$	166.08	0	0.42486	1	3
$\varphi(t)pent(t)p(t)$	168.16	2.08	0.15027	0.3537	4
$\varphi(.)pent(.)p(t)$	59493.78	59327.7	0	0	2
$\varphi(t)pent(.)p(.)$	59510.56	59344.48	0	0	2
$\varphi(t)pent(.)p(t)$	59512.61	59346.54	0	0	3
$\varphi(.)pent(t)p(.)$	59535.05	59368.97	0	0	2
$\varphi(.)pent(.)p(.)$	59538.06	59371.98	0	0	3

* φ , survival probability; p , capture probability; $pent$, recapture probability; $(.)$, constancy; (t) , temporal variation

Appendix B. Size Classes, by Year (Alternative)

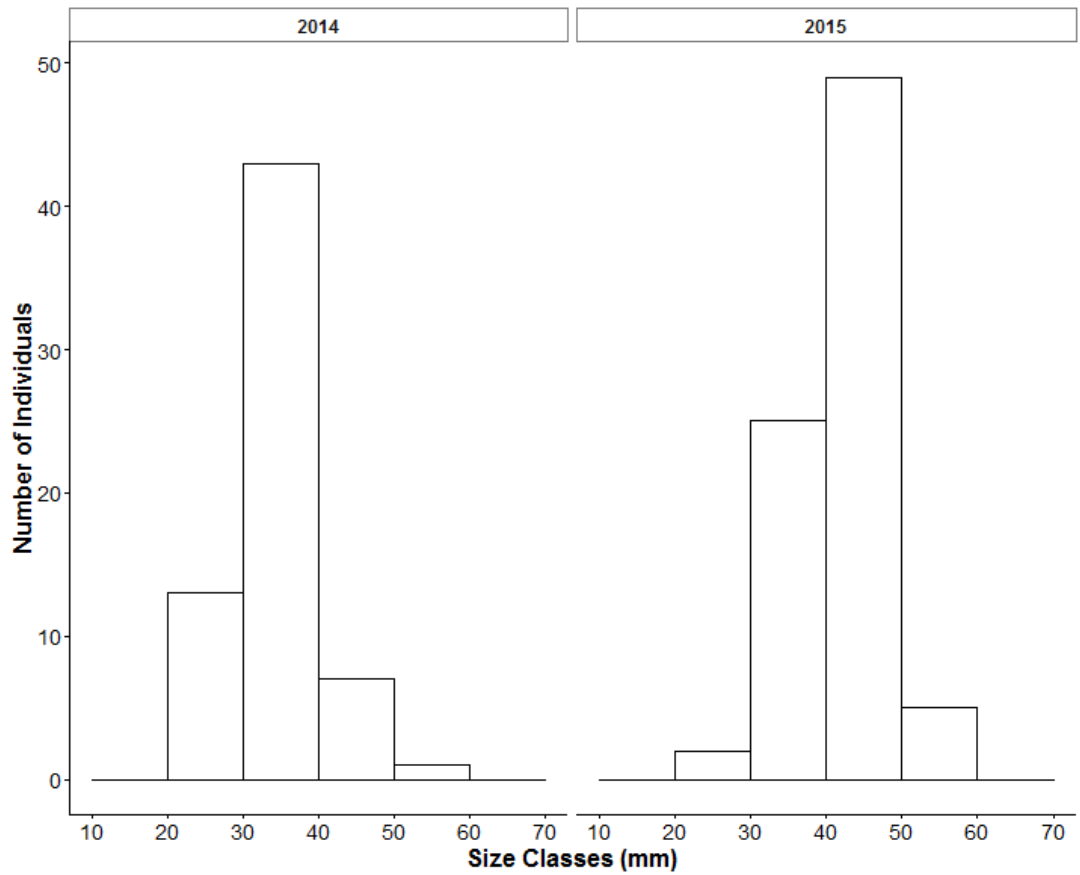


Figure B.1 Sizes classes for *Pleurobema riddellii* at *Neches 1* site.

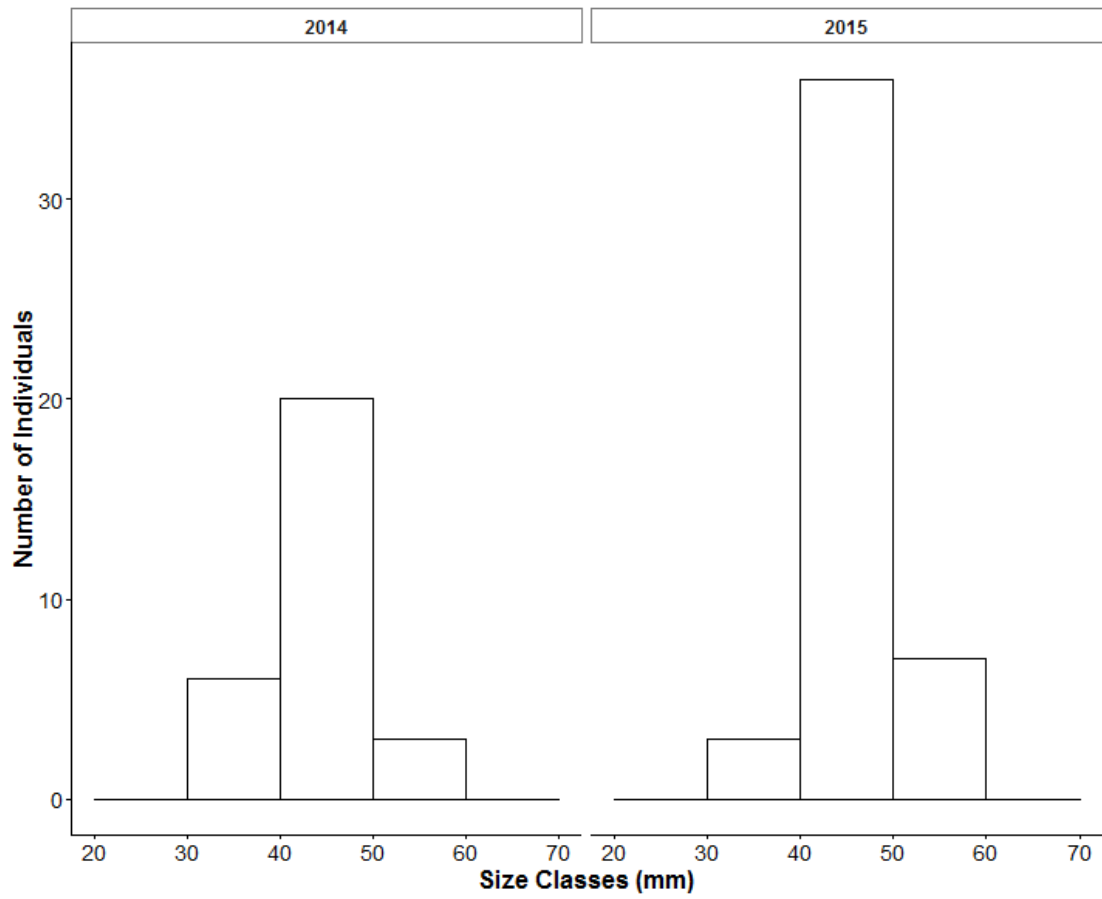


Figure B.2 Sizes classes for *Pleurobema riddellii* at Neches 2 site.

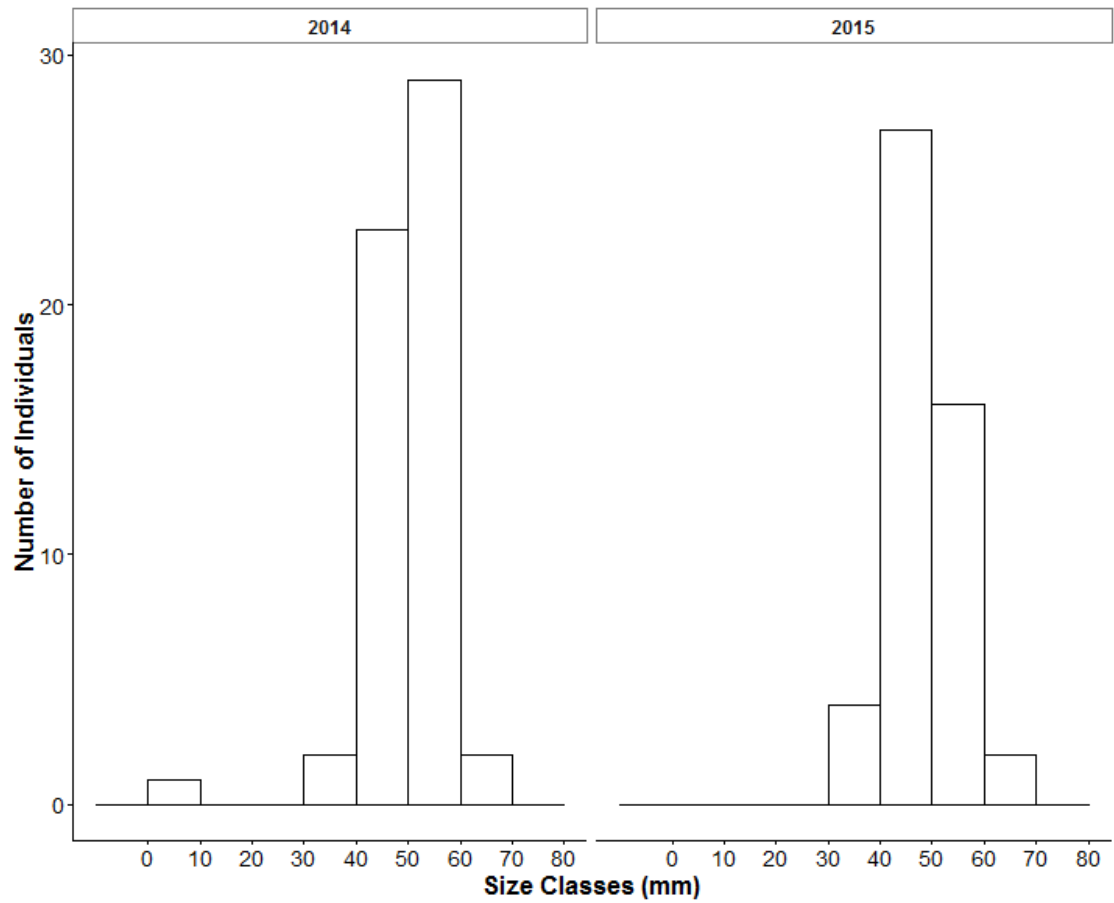


Figure B.3 Sizes classes for *Pleurobema riddellii* at Neches 3 site.

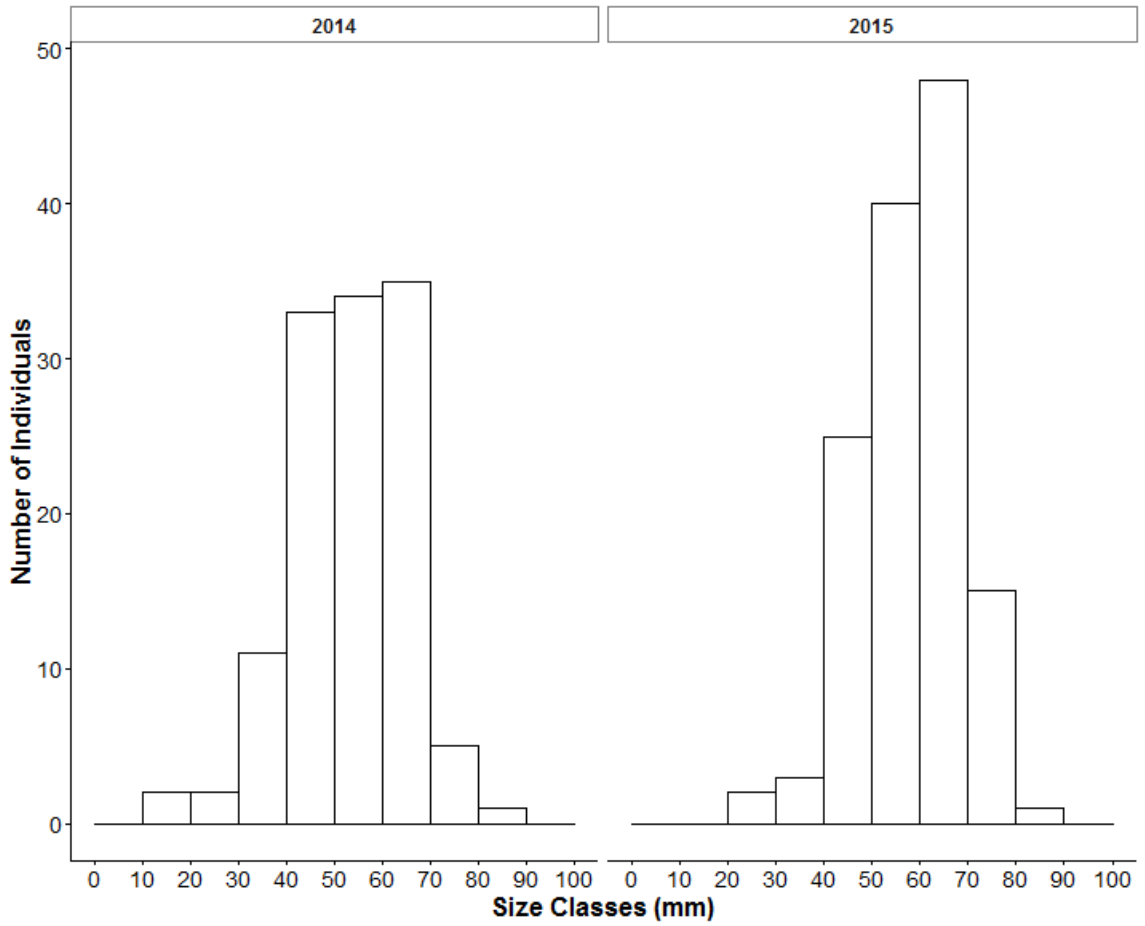


Figure B.4 Sizes classes for *Fusconaia askewi* at Sabine I site.

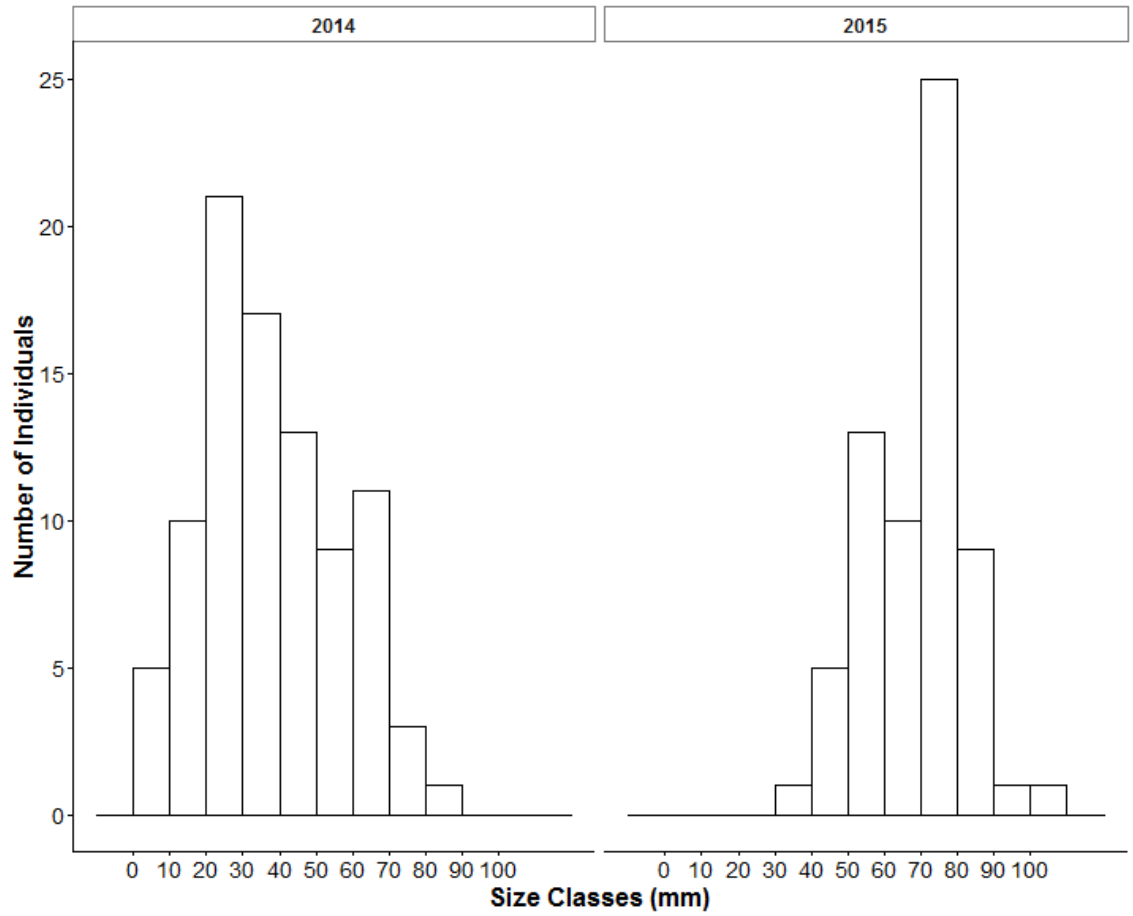


Figure B.5 Sizes classes for *Fusconaia askewi* at Sabine 2 site.

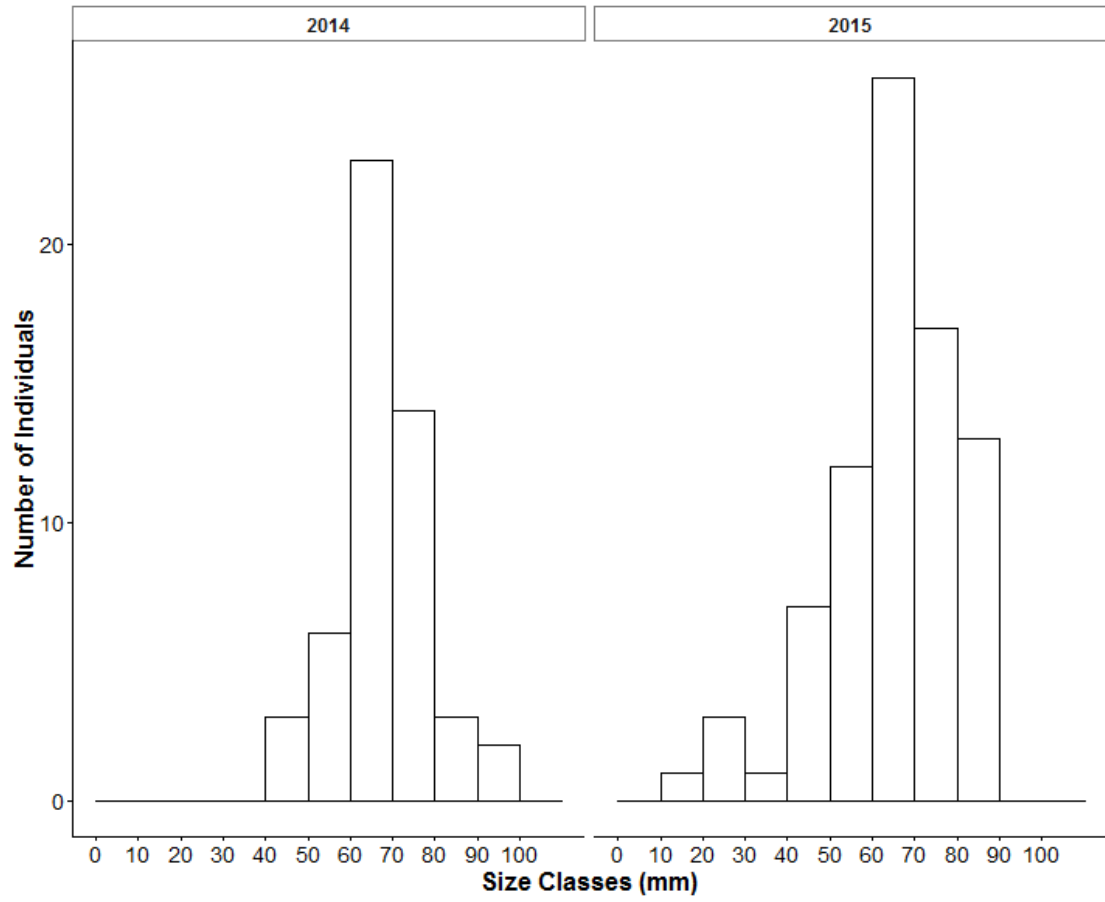


Figure B.6 Sizes classes for *Fusconaia askewi* at Sabine 3 site.

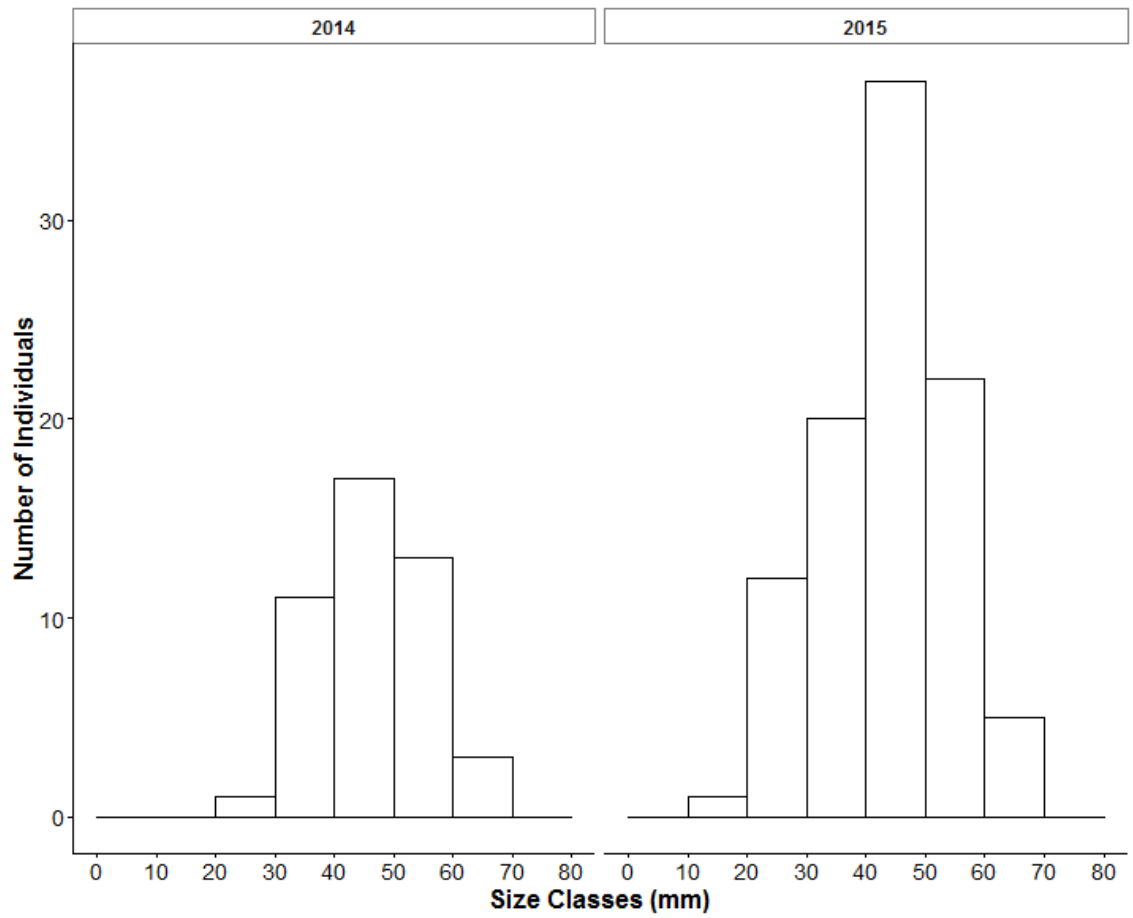


Figure B.7 Sizes classes for *Fusconaia askewi* at Angelina 1 site.

Appendix C: Nucleotide alignment of the *ND1* gene

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Faskewi EP138      ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Faskewi EP134      ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Faskewi EP139      ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Faskewi EP136      ACATAACCTCCACACTTATTACATATCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Faskewi EP130      ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP146  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP151  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP107  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP154  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP152  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP153  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP147  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP145  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
Flananensis EP150  ACATAACCTCCACACTTATTACATACCTTCTAATCTTACTAGGCGTAGCATTCTTTACCC
GENBANK Fflava
AY613793

Faskewi EP138      TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Faskewi EP134      TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Faskewi EP139      TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Faskewi EP136      TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Faskewi EP130      TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP146  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP151  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP107  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP154  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP152  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP153  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP147  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP145  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
Flananensis EP150  TTCTTGAACGCAAAGCTTTAGGGTACTTTCAAATCCGAAAAGGCCCAAACAAAGTTGGAA
GENBANK Fflava
AY613793
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Appendix C (Continued)

Faskewi EP138 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Faskewi EP134 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Faskewi EP139 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Faskewi EP136 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Faskewi EP130 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP146 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP151 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP107 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP154 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP152 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP153 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP147 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP145 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
Flananensis EP150 TTATAGGAATCCCACAACCACTAGCAGACGCCCTAAAAC TTTTGTGAAAGAATGAGTAA
GENBANK Fflava
AY613793

Faskewi EP138 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Faskewi EP134 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Faskewi EP139 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Faskewi EP136 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Faskewi EP130 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP146 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP151 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP107 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP154 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP152 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP153 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP147 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP145 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
Flananensis EP150 TACCACATCCTCAAAC TACTTACCATTTATTTTAACCCCAACAATCATATTAATTTTAG
GENBANK Fflava
AY613793

Appendix C (Continued)

Faskewi EP138 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Faskewi EP134 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Faskewi EP139 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Faskewi EP136 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Faskewi EP130 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP146 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP151 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP107 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP154 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP152 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP153 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP147 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP145 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
Flananensis EP150 CACTTAGGCTATGACAGCTATTCCCATCCTTTATACTCTCATTTCAAATAACCCTAGGAA
GENBANK Fflava
AY613793 CACTTAGACTATGACAACCTATTCCATCCTTTATACTCTCATTTCAAATAGCCCTAGGAA

Faskewi EP138 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Faskewi EP134 TACTCATATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Faskewi EP139 TACTCATATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Faskewi EP136 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Faskewi EP130 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP146 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP151 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP107 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP154 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP152 TACTCATATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP153 TACTCATATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP147 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP145 TACTCATATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
Flananensis EP150 TACTCCTATTCTTATGTATTTCTTCTTTAACCGTCTACACAACCTTAATAGCAGGTTGGG
GENBANK Fflava
AY613793 TACTCTTATTCTTATGTATCTCTCCTTAAGTCTATACAACCTTAATAGCAGGTTGGG

Appendix C (Continued)

Faskewi EP138 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Faskewi EP134 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Faskewi EP139 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Faskewi EP136 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Faskewi EP130 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP146 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP151 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP107 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP154 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP152 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP153 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP147 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP145 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
Flananensis EP150 CCTCAAACCTCGAAGTATGCTCTACTAGGAGCCATTCGAGCCATGGCCCAAACCATCTCAT
GENBANK Fflava
AY613793

Faskewi EP138 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Faskewi EP134 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Faskewi EP139 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Faskewi EP136 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Faskewi EP130 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP146 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP151 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP107 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP154 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP152 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP153 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP147 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP145 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
Flananensis EP150 ATGAAGTAACAATAACACTAATTATCATTCTTCTACCTATTCTTGATTATACAAATAGACA
GENBANK Fflava
AY613793

Appendix C (Continued)

Faskewi EP138 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Faskewi EP134 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Faskewi EP139 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Faskewi EP136 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Faskewi EP130 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP146 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP151 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP107 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP154 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP152 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP153 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP147 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP145 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
Flananensis EP150 TAGTAACAATCCGCTCGGTAAACACCTCTATACCAACCTTTGCCCTCTCCGCACCATTAG
GENBANK Fflava
AY613793

Faskewi EP138 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Faskewi EP134 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Faskewi EP139 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Faskewi EP136 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Faskewi EP130 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP146 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP151 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP107 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP154 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP152 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP153 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP147 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP145 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
Flananensis EP150 CCATTATGTGAAGTGTGTCATCTTAGCAGAAACAAACCGAGCCCCATTTGACTTTGCTG
GENBANK Fflava
AY613793

Appendix C (Continued)

Faskewi EP138 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Faskewi EP134 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Faskewi EP139 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Faskewi EP136 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Faskewi EP130 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP146 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP151 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP107 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP154 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP152 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP153 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP147 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP145 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
Flananensis EP150 AAGGGGAATCAGAACTAGTCTCTGGATTTAATATTGAGTACGGCGGAGCCGGCTTTGCTT
GENBANK Fflava
AY613793

Faskewi EP138 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Faskewi EP134 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Faskewi EP139 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Faskewi EP136 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Faskewi EP130 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP146 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP151 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP107 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP154 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP152 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP153 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP147 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP145 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
Flananensis EP150 TCCTCTTTATAGCCGAATACAGTAACATCTTAATAATAAGACTCCTTACTGCCTGTATAC
GENBANK Fflava
AY613793

Appendix D: Nucleotide alignment of the *COX1* gene

FaskewiGENBANK JN180998	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FaskewiGENBANK KT285626	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP107	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP150	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP151	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP147	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP106	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP146	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FlananensisEP145	CTTTATGATCTGGTTTGATTGGGTTGGCTTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FflavaGENBANK KT285636	CTTTATGATCTGGTTTGATTGGATTGGCTTAAGTCTTTTGATTTCGAGCTGAGTTAGGGC
FaskewiGENBANK JN180998	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FaskewiGENBANK KT285626	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP107	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP150	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP151	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP147	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP106	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP146	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FlananensisEP145	AGCCAGGAAGGTTGTTGGGGGATGATCAGTTGTATAAATGTGATTGTGACGGCGCATGCTT
FflavaGENBANK KT285636	AGCCCGGTAGGTTGTTGGGGGATGATCAATTGTATAAATGTGATTGTGACGGCGCATGCTT

Appendix D (Continued)

FaskewiGENBANK JN180998	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FaskewiGENBANK KT285626	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGTGGTTTTGGTAATTGGCTTA
FlananensisEP107	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP150	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP151	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP147	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP106	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP146	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FlananensisEP145	TTATAATAATTTTCTTTTGGTGATACCTATGATGATTGGAGGTTTTGGTAATTGGCTTA
FflavaGENBANK KT285636	TTATAATAATTTTCTTTTGGTGATACCTATGATAATTGGTGGTTTTGGTAATTGGCTTA
FaskewiGENBANK JN180998	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FaskewiGENBANK KT285626	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP107	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP150	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP151	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP147	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP106	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP146	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGATTAATAATCTAAGGTTTT
FlananensisEP145	TTCTCTTATGATTGGGGCTCCGGATATGGCTTTTCTCGGTTAATAATCTAAGGTTTT
FflavaGENBANK KT285636	TTCTCTTATGATTGGAGCTCCTGATATGGCTTTTCTCGATTGAATAATTTGAGGTTTT

Appendix D (Continued)

FaskewiGENBANK JN180998	GGT-TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTAGTGGAGAGGGGTGT
FaskewiGENBANK KT285626	GGT-TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FlananensisEP107	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FlananensisEP150	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FlananensisEP151	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FlananensisEP147	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FlananensisEP106	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTAGTGGAGAGGGGTGT
FlananensisEP146	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTAGTGGAGAGGGGTGT
FlananensisEP145	GGT--TACTTGTGCCTGCTCTTTTTTTATTGCTAAGATCTTCTTTGGTGGAGAGGGGTGT
FflavaGENBANK KT285636	GGT-TACTTGTGCCCTGCTCTTTTTTTGTTAAGATCTTCTTTGGTGGAGAGGGGTGT
FaskewiGENBANK JN180998	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FaskewiGENBANK KT285626	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP107	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP150	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP151	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP147	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP106	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP146	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FlananensisEP145	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAATATTGCTCATTCTGGAGCTTC
FflavaGENBANK KT285636	TGGGACTGGTTGAACGGTTTATCCGCCGTTGTCTGGGAACATTGCTCATTCTGGAGCTTC

Appendix D (Continued)

FaskewiGENBANK JN180998	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FaskewiGENBANK KT285626	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FlananensisEP107	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FlananensisEP150	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCCAT
FlananensisEP151	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCCAT
FlananensisEP147	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FlananensisEP106	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FlananensisEP146	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FlananensisEP145	AGTGGATTTGGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATTTTGGGGGCTAT
FflavaGENBANK KT285636	AGTGGATTTAGCTATTTTTCTTTGCATCTTGCTGGTGCATCTTCTATCTTGGGGGCTAT
FaskewiGENBANK JN180998	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FaskewiGENBANK KT285626	TAAC TTTATTTCTACTGTAGGCAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP107	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP150	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP151	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP147	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP106	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP146	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FlananensisEP145	TAAC TTTATTTCTACTGTAGGTAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC
FflavaGENBANK KT285636	TAAC TTTATTTCTACTGTGGGAATATGCGGTCTCCAGGATTGGTTGCTGAGCGAATTCC

Appendix D (Continued)

FaskewiGENBANK	GTTATTCGTGTGGGC-----
JN180998	
FaskewiGENBANK	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
KT285626	
FlananensisEP107	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP150	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP151	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP147	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP106	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP146	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FlananensisEP145	GTTATTCGTGTGGGCTGTAACGGTAACAGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
FflavaGENBANK	GTTGTTTCGTGTGGGCTGTAACGGTAACGGCGGTTTTGTTGGTTGCTGCGTTGCCTGTTTT
KT285636	
FaskewiGENBANK	-----
JN180998	
FaskewiGENBANK	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
KT285626	
FlananensisEP107	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP150	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP151	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP147	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP106	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP146	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FlananensisEP145	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
FflavaGENBANK	AGCTGGTGCC-ATTACGATGTTGCTTACTGATCGTAATATTAACACGTCTTTTTTTGATC
KT285636	

Appendix E: Nucleotide alignment of the *16S* gene

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EP130          AATGCCTGCCAGTGAAAACTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP136          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP138          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP139          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP140          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP145          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP150          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP151          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP152          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
EP153          AATGCCTGCCAGTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
GenBank_      -----GTGAAAATTTTTAAACGGCCGCGTTAGCGTGAGCGTGCTAAGGTAGCG
AY238481_F_flava
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EP130          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP136          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP138          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP139          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP140          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP145          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP150          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP151          TTTTAAATAGCCTTTTAAATTGGGGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP152          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
EP153          TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCTGTACCTT
GenBank_      TAATAAATAGCCTTTTAAATTGGAGGCCAGTGAATGGCAAGACTAGGAATACCT-TACCTT
AY238481_F_flava
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Appendix E (Continued)

EP130 TTATGAAAAAAAAAAGCTTTTCACCTGAGTGAAAAGACTCAGATAGCAAAGGAAGACGAAAA
EP136 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP138 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP139 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP140 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP145 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP150 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP151 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP152 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
EP153 TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGATAGCGAAGGAAGACGAAAA
GenBank_
AY238481_F_flava TTATGAAAAAAAAAAGCTTTTCATCTGAGTGAAAAGACTCAGA-AGCGAAGGAAGACGAAAA

EP130 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCACAAACACAAAAGACAAAAGG
EP136 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP138 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP139 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP140 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP145 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP150 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP151 GAGCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCCTTAAACACAAAAGACAAAAGG
EP152 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
EP153 GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCATAAACACAAAAGACAAAAGG
GenBank_
AY238481_F_flava GACCCCGCGGAAGCTTTACCTTTTCCAGCCTTAGCTGCCACAAACACAAAAGGCAAAAAGG

Appendix E (Continued)

EP130 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAATATTCAT
EP136 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP138 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP139 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP140 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP145 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP150 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP151 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP152 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
EP153 TTTGATTGGGGCAATCTCGGAACAACCAAGCTTCCGATTCTCCTTAAAT-AAACATCCAT
GenBank_
AY238481_F_flava TTTGAT-GGGGCAATCTCGGAACAACCAAGCTTCCGATTCTACTTAAGTGAACATCCAT

EP130 AACCTGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP136 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP138 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP139 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP140 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP145 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP150 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP151 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP152 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
EP153 AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATAACAGCGTAATCCAGCTCAAGAG
GenBank_
AY238481_F_flava AACCCGATAAGGACAAAAAGAAGTTACCCCGGGGATA-CAGCGTAATCCAGCTCAAGAG

Appendix E (Continued)

EP130 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATTA
EP136 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP138 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP139 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP140 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP145 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP150 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP151 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP152 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
EP153 -TACACATCGAAAGCTGGGTTTGCG-ACCTCGATGTTG--GCTTAAGGACATCCACATCA
GenBank_AY238481_F_flava CTACACATCGAAAGCTGGGTTTGCGCACCTCGATGTTGCCGCTTAAGGACATCCAC----

Appendix F: Nucleotide alignment of the *ITS1* gene

EP130 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP134 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP135 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP136 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP139 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP145 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP146 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP150 GATCATTTCCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP152 GATCATTTCCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP153 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP107 GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
EP106 GATCATTTCCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
Fusconiaflava GATCATTACCGAAAAGATTTAAAAAGCGCTGGCCGTTTTTTAGATCGAGAGACACAAAAG
ITS_DQ383442

EP130 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP134 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP135 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP136 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP139 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP145 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP146 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP150 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP152 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP153 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP107 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
EP106 CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
Fusconiaflava CGCATCCGAAAACGAGAAAAAGACTGGGTTGCGGAGGTGGCTAGATCGCCTCCTTCCCG
ITS_DQ383442

Appendix F (Continued)

EP130 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP134 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP135 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP136 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP139 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP145 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP146 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP150 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP152 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP153 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP107 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
EP106 GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
Fusconiaflava GTCTATAAACCTGTGTAGATCCATGGCCGCCGGTCGGGGGTACCTAAGTCCGAAGTAGGC
ITS_DQ383442

EP130 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP134 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP135 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP136 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP139 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP145 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP146 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP150 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP152 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP153 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP107 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
EP106 CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
Fusconiaflava CCGCAATGCTTCAAGCGGGGCATGTTCGACGAGAGCTTTGGTACAGCGATGGTTTAAAGA
ITS_DQ383442

Appendix F (Continued)

EP130 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP134 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP135 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP136 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP139 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP145 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP146 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP150 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP152 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP153 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP107 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
EP106 GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
Fusconiaflava GAGAGCTTCCAGCTACGCCCCGCCACACATTCTTTCCGCCCCGTGCGGTTTTCTCTTT
ITS_DQ383442

EP130 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP134 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP135 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP136 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP139 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP145 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP146 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP150 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP152 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP153 GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP107 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
EP106 GCGTTTGACACTTGTAACCAAAAGACGGCGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
Fusconiaflava GCGTTTGACACTTGTAACCAAAAGTCGACGGCTCGATTTGGCCAGGTTGGCTCCGTTTCT
ITS_DQ383442

Appendix F (Continued)

EP130 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP134 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP135 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP136 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP139 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP145 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP146 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP150 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP152 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP153 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP107 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
EP106 CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
Fusconiaflava CTTGTCCATTTTGGGCCCCGTCGATGCAAGGGATCGATCCGTCTCGTTCGTTGCCGGACG
ITS_DQ383442

EP130 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP134 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP135 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP136 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP139 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP145 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP146 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP150 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP152 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP153 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP107 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
EP106 GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
Fusconiaflava GCCAATCGATCCTAATCCTTGCAAGCTCCCAGCTCGGGGGCGAGAAAGGAAATTACTCT
ITS_DQ383442