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Physiological modulation of the vitality of Scots pine trees by atmospheric ammonia deposition

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2. Foliar uptake of gaseous ammonia²

2.1. Abstract

Young saplings of Pinus sylvestris L. were exposed to gaseous NH₃ at 53 or 105 µg m⁻³ for one year in opentop chambers. Saplings received ¹⁵Nlabelled $(NH_4)_2SO_4$ via the soil. To examine the importance of foliar N uptake, changes in the concentration of total and labelled N in the needles were followed. Increases in needle biomass and N concentration were found in trees exposed to NH₃, confirming that atmospheric NH₃ acted as a N fertilizer. NH₃ had a greater and quicker effect than $(NH_4)_2SO_4$: compared with the growth in amb ient air, the N concentration in the needles exposed to NH₃ had increased by 49% in four months, while the increase after the highest N-fertilization (200 kg N ha $^{-1}$ y⁻¹) was only 8%. The small contribution of NH4⁺ fertilization to the total N concentration was not due to a deficient N uptake: the ¹⁵N concentration in the needles increased significant ly with time. On the other hand, NH 3 uptake in shoots may have had a negative effect on the NH_4^+ root uptake. The relation between plant N and atmospheric NH₃ concentration was non-linear and possible reasons for this observation are discussed. Fumigation with NH₃ significantly

decreased the ratios of K/N and P/N, showing that fumigation disrupted the nutrient ba lance.

2.2. Introduction

Deposition of atmospheric N has increased during the last decades, especially in the Netherlands (Asman et al., 1988; Maas, 1991). It is probably an important cause of forest decline in Europe. Conifers and hardwoods planted on soils originally poor in available N, are in the most difficult situation (Van Breemen & Van Dijk, 1988). The NH_v deposition (NH₃ + NH₄⁺) on the forests in the Netherlands is higher than average over the country as a whole: 80 compared with 47 kg N $ha^{-1} y^{-1}$ (Houdijk & Roelofs, 1991) and N deposition is assumed to be implicated in the damage to forest ecosystems (Boxman et al., 1986; Heinsdorf, 1991; Nihlgård, 1985; Roelofs et al., 1985; Van Dijk et al., 1990). Scots pine (Pinus sylvestris L.) is the most common species in boreal forests and its vitality is low nearly all over Central and Western Europe. Heij & Schneider (1991), Horn et al. (1989), Temmerman &

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Coosemans (1988) and Van Dijk & Roelofs (1988) found that Scots pine with yellow ne edles have a very high N concentration of 2.8 to 3% dry weight, far above the 1.8% considered optimal for production (Von Heinsdorf & Krauß, 1991). They concluded that this is caused by excess atmospheric N entering the tree through roots and shoots.

Hutchinson et al. (1972) and Porter et al. (1972) found that leaves are efficient sinks for gase ous NH₃. Later reports (Dueck et al., 1991; Lalisse & Bardin, 1981; Van der Eerden et al., 1990; Whitehead & Lockyer, 1987) confirmed that atmospheric NH₃ can be used by the plant as a source of N. Ammonia can be absorbed through the cuticle but it mainly enters the plant via the stomata (Hosker & Lindberg, 1982). Van Hove et al. (1989) found that only 3% of the NH₃ absorbed by leaves of Phaseolus vulgaris L. entered via the cuticle while the remainder passed via the sto mata. The dominant role of stomata in NH₃ absorption was also found in con ifers (Van Hove et al., 1992). N absorption takes place in two phases: first it is stored and then it is m etabolized (Lalisse & Bardin, 1981), in which NH₃ is incorporated into organic N via the gluta mine synthetase-glutamate synthase pathway (Miflin & Lea, 1976). These organic N compounds are partly transported from the needles via the phloem to the roots where they can meet the N requirement (Lemon & Van Houtte, 1980; Rowland, 1986; Rowland et al., 1987).

Cycling-organic N compounds play a major role in regulation of nitrate uptake (Cooper et al., 1986; Rowland et al., 1987). We suggest that direct foliar uptake of NH₃ will increase the N concentration in the needles and transport to the roots where it will reduce the N uptake.

The object of our experiment was to test this hypothesis: the i ncorporation of needle N from NH 3 fumigation was measured as well as by uptake fertilizing Ν with $(NH_4)_2SO_4$. We chose NH_4^+ instead of NO_3^- as the N source, because conifer seedlings and mycorrhizas show preference for NH_4^+ (Boxman & Roelofs, 1988; Finlay et al., 1989a; Kamminga-van Wijk, 1991; Zedler et al., 1986). SO_4^{2-} was chosen because this is an important anion in precipitation (Heij & Schneider, 1991). Scots pine trees were exposed to gaseous NH₃ in open-top chambers and received ¹⁵Nlabelled $(NH_4)_2SO_4$ via the soil. The changes in N concentrations of the needles were followed for a year so that seasonal changes in N uptake in relation to the N requirement for growth were recorded. In addition, we studied the influence of NH₃ on the nutrient balance and water d emand implied by earlier experi ments (Van der Eerden & Pérez-Soba, 1992; Van der Eerden et al., 1992) and the nutrient composition and water potential of the needles were examined.

2.3. Materials and methods

Table 2.1. Design of the experiment in which Pinus sylvestris L. trees were exposed to ambient air (AA) and AA supplemented with 46 or 98 μ g m⁻³ NH₃ and fertilized with (NH₄)₂SO₄ (enriched with ¹⁵N) over one year.

NH ₃ fumigation $(\mu g m^{-3})$	(Nl tio	$H_4)_2SO_4$ fertiliza- n (kg N ha ⁻¹ y ⁻¹)
7 (AA)	0	50 100 200
53	0	50
105	0	50

Plant material and soil type

Two-year-old Scots pine trees (Pinus sylvestris L.) of the same provenance were potted in nutrient-poor acidic sandy soil in 10 l pots. The soil had a pH(CaCl₂) of 5.2 and a nutrient composition of 0.20 % dry weight total N, 6.94 μ g g⁻¹ NH₄⁺, 9.43 μ g g⁻¹ NO₃⁻, 1.31 μ g g⁻¹ P and 24.9 μ g g⁻¹ K. After planting, the trees were kept outdoors for 11 months until the start of the experiment, which ran from August 1988 till August 1989.

Fumigation and fertilization

In August 1988, we chose seventytwo trees of similar height and with about the same number of branches. We divided them at random into eight groups, each with nine repl icates, and placed them in open-top chambers. For one year, four of the groups were exposed to ambient air (AA) and the other four to AA su pplemented with 46 or 98 μ g m⁻³ NH₃. The annual mean concentra tion of NH₃ in AA was 7 μ g m⁻³. A description of the chambers and NH₃ fumigation is given by Dueck (1990). Table 2.1 is a summary of the experimental design.

Once a month, a (NH₄)₂SO₄ solution enriched in ¹⁵N, (¹⁵N abundance of 2.260 atom % compared with a natural abundance of 0.366 atom %) was applied as a fertilizer to the soil, at N concentrations equivalent to 50, 100 and 200 kg N ha⁻¹ y⁻¹ (hereafter referred to as kg N) to a number of the groups exposed to AA or to 53 and 105 μ g m⁻³ NH₃ (Tab. 2.1). We chose a fertilization of 50 kg N because it is close to the ave rage deposition in the Netherlands. The ^{15}N enriched (NH₄)₂SO₄ solution was supplied to field capacity to prevent loss of fertilizer. Hence, there were no losses of seepage w ater. Since other vegetation competes strongly for the fertilizer, we weeded once in the winter and once a month from April onwards. We watered the trees when necessary to prevent a low soil water status that could make the fertilization useless or even harmful (Van Breemen & Van Dijk, 1988).

Assessment of parameters in the needles

At the end of the experiment we harvested the needles and sorted them into current-year and one-year-old needles. After oven-drying them at 75°C for 24 h, we weighed and ground them, and determined total N (Carlo Erba Elemental Analyzer). We also dry-ashed samples at 600 °C and measured the concentrations of P, Ca and Mg, using a Technicon Auto Analyzer and K by flame spe c-

trophotometry. We determined the ¹⁵N atom % abundance by mass spectrometry using the method d escribed by Dueck et al. (1991). To determine the seasonal variations in concentration and dry-weight Ν percentage (% fresh weight) in the needles, five times during the e xperiment we took samples of approx. 0.5 g fresh weight. Samples were taken from the central branch at the top of the crown because the N co ncentration in the needles signif icantly varies according to their pos ition in the crown (Evers et al., 1991). We measured the current-year and one-year-old top shoots at maximum length.

We derived the following formula to estimate the respective contributions of NH $_3$ fumigation and $(NH_4)_2SO_4$ fertilization to the needle N concentration:

$$\Delta N = \Delta N_{\text{NH}_3} + \Delta N_{(\text{NH}_4)_2} SO_4$$
(1)

where:

 ΔN = increase in N concentr ation (% dry-weight) due to the NH _y treatment, calculated by subtracting the N concentration in the needles of trees exposed to AA from the N concentration in the needles exposed to AA supplemented with NH ₃ and fertilized with (NH ₄)₂SO₄.

 ΔN_{NH_3} = increase in N conce ntration due to gaseous NH ₃

 $\Delta N_{(NH_4)_2SO_4}$ = increase in N concentration due to uptake by the roots from the NH₄⁺ fertilizer and transport to the needles, calculated as follows:

Table 2.2. Needle biomass (g DW) and length increment of the current-year central shoot with respect to the previous year (% previous year length) of Pinus sylvestris L. exposed to ambient air (AA), 53 and 105 μ g m⁻³ NH₃ and fertilized with (NH₄)₂SO₄ (kg N ha⁻¹ y⁻¹). Means (n = 9) in each column with the same letter are not significantly different at α = 0.05.

NH ₃ fumigation	(NH ₄) ₂ SO ₄ fertilization	Needle bior	nass	Shoot length increment				
		Current year	One year-old	_				
AA	0	20.7 ^a	11.3 ^a	38.6 ^a				
AA	50	18.3 ^a	10.8 ^{ab}	58.7 ^a				
AA	100	20.1 ^a	10.4 ^{ab}	69.8 ^a				
AA	200	28.6 ^b	10.7^{ab}	82.8 ^a				
53	0	31.8 ^b	6.8 ^{bc}	64.3 ^a				
53	50	31.3 ^b	4.8 ^c	67.4 ^a				
105	0	32.8 ^b	5.9 ^c	57.4 ^a				
105	50	34.1 ^b	$8.7^{\rm abc}$	56.8 ^a				

$$\frac{a-b}{c-b} \times (\% \text{ N needles})$$
(2)

where a, b and c are, respectively, the atom % ¹⁵N in the needle sample, in the air and in the added NH₄⁺. We made the following assumptions in the calculation of (2): ¹⁴N and ¹⁵N behave similarly in the tree, the nat ural difference in ¹⁵N-percentage b etween AA and needles was neglig ible (Vose, 1980) and the needles did not loose N. By substitution of (2) in (1) ΔN_{NH_3} can be calculated. The incorporation rate of total N into the needles by NH₄⁺ fertilization was calculated from (2).

Water potential

We measured the water potential in the needles using the pressure cha mber described by Scholander et al. (1965).

Statistics

We evaluated needle biomass, the concentration of N, Ca, Mg, P, K, ¹⁵N and the uptake rate of total N from the fertilizer with analyses of variance. We performed a square root transformation to stabilize the variance of the ¹⁵N content and the data on the rate of uptake of total N from the fertilizer. After the analyses of variance we used an LSD-test for testing pairwise diffe rences.

Water potential data were studied by analyses of variance a ccording to a split-plot design, with

Figure 2.1. Changes in total N concentration (% DW) with time in needles of Pinus sylvestris L. exposed to ambient air and fertilized with $(NH_4)_2SO_4$ (kg N ha⁻¹ y⁻¹). Each bar diagram represents the mean of nine measurements and the vertical lines represent the LSD at $\alpha = 0.05$ for each harvest.

fertilization on whole plots and time on subplots.

2.4. Results

Shoot growth

We measured growth in terms of needle biomass production and increase in length of the central shoot compared to that in the previous year (Table 2.2). Data concern the bi omass of the remaining needles. The biomass of previous year's needles (one-year-old) decreased signif icantly after NH₃ fumigation (loss of needles), while the production of current-year needles showed a si gnificant increase after NH₃ fumigation and after the largest fertiliz ation, 200 kg N. The increase in length of the central shoot was greater with respect to that of the previous year after fumigation and fertilization, although differences were not si g-nificant.

There was a general increase in the dry weight p ercentage with time

Figure 2.2. Changes in total N concentration (% DW) with time in needles of Pinus sylvestris exposed to ambient air, to 53 and to 105 μ g m⁻³ NH₃ and not fertilized (a) or fertilized with (NH₄)₂SO₄ to 50 kg N ha⁻¹ y⁻¹ (b). Each bar diagram represents the mean of nine measurements and the vertical lines represent the LSD at $\alpha = 0.05$ for each harvest.

Table 2.3. ¹⁵N concentration (atom % excess) and uptake rate of total N (mg N g⁻¹ needle FW month⁻¹) by (¹⁵NH₄)₂SO₄ fertilization (1.894 atom % excess ¹⁵N) into one-year-old needles of Pinus sylvestris exposed to ambient air (AA), to 53 and to 105 μ g m⁻³ NH₃ and fertilized with (¹⁵NH₄)₂SO₄. Contrast between different fumigation (AA, 53 and 105 μ g m⁻³ NH₃) and fertilization (50,100 and 200 kg N ha⁻¹ y⁻¹) treatments, and their interactions with time were tested by analyses of variance. *, P<0.05; **, P<0.01; ***, P<0.001; ns, not significant.

NH ₃ fumigation	(NH ₄) ₂ SO ₄ fertilization	¹⁵ N	¹⁵ N N uptake rate								
		Dec	Feb	May	Aug	Dec	Feb	May	Aug		
AA	50	0.006	0.020	0.057	0.063	0.009	0.021	0.039	0.027		
AA	100	0.007	0.051	0.108	0.145	0.013	3 0.058	0.069	0.051		
AA	200	0.068	0.190	0.371	0.291	0.144	0.237	0.278	0.155		
53	50	0.004	0.017	0.044	0.044	0.008	3 0.027	0.044	0.027		
105	50	0.006	0.019	0.043	0.040	0.015	5 0.039	0.056	0.025		
Analyses of variance											
Time		***				***					
NH ₃ fumigation		ns					ns				
(NH ₄) ₂ SO ₄ fertil	ization	***	***				***				
Time x fumigati	on	ns				ns					
Time x fertilizat	tion	*				**					

from August 1988 to August 1989 (no data given), that was only inte rrupted in the spring. In August 1989, the dry weight percentage of the oneyear-old needles was significantly higher in the fumi gated trees than in the non-fumigated ones, while in the current-year needles, the dry weight percentage was significantly lower in the fumigated trees.

Changes in total N concentration in needles

The N concentration of the one-yearold needles was not significantly affected by $(NH_4)_2SO_4$ fertilization, except with 200 kg N in August 1989 (Fig. 2.1). It increased very significantly (P < 0.001) with NH₃ fumigation, both with or without 50 kg N (Figs. 2.2a and 2.2b). Interes tingly, the N concentration of oneyear-old needles of unfertilized trees was not much different for the two NH₃ fumigation treatments; for fe rtilized trees, the higher NH₃ fumigation resulted in a higher N conce ntration from August 1988 till May 1989. No significant interaction was found between NH₃ fumigation and $(NH_4)_2SO_4$ fertilization. The tendency found for current-year needles was the same as found for one-year-old needles, a 1

Table 2.4. ¹⁵ N concentration (atom % excess) and uptake rate of total N (mg N g ⁻¹ needle FW month ⁻¹) by (NH ₄) ₂ SO ₄
fertilization (1.894 atom % excess ¹⁵ N)
into current-year needles of Pinus
sylvestris exposed to ambient air (AA),
to 53 and to 105 μ g m ⁻³ NH ₃ and fertil-
ized with $(NH_4)_2SO_4$. Needles emerged
in May 1989. Contrast between different
fumigation (ambient, 53 and 105 μ g m ⁻³
NH ₃) and fertilization (50, 100 and 200
kg N ha ⁻¹ y ⁻¹) treatments was tested by analyses of variance. ***, $P<0.001$.
,1 0.001.

NH ₃ fumigation	(NH ₄) ₂ SO ₄ fertilization		N uptake rate							
AA	50	0.160	0.243							
AA	100	0.276	0.386							
AA	200	0.543	0.877							
53	50	0.094	0.153							
105	50	0.095	0.170							
Analyses of v	Analyses of variance									
NH ₃ fumigati	on	***	***							
(NH ₄) ₂ SO ₄ fe	ertilization	***	***							

though its statistical significance was less.

Changes in ¹⁵N concentration in needles

Data on one-year-old and currentyear needles are shown in tables 2.3 and 2.4, respectively. In one-yearold needles, ¹⁵N concentration inwith time in creased all the $(NH_4)_2SO_4$ fertilization treatments. This increase was more than propo rtional: the results suggest a curve of the second order in which the ${}^{15}N$ concentration proportionally i ncreased with the square of the exp osure time.

The contribution of 50 kg N to the ¹⁵N concentration did not differ significantly between fumigated and non-fumigated trees, although it was lower in the fumigated trees. There was a very clear response to NH $_4^+$ fertilization, with the highest ^{15}N concentration observed with the highest concentration of fertilizer. We found a significant interaction between time and fertilization. Cu rrent-year needles, formed during the experimental period, had a significantly higher (P<0.001) ¹⁵N concen-

Table 2.5. Contribution (%) of NH₃ fumigation (FU) and $({}^{15}NH_4)_2SO_4$ fertilization (FE) to the increase in N concentration in needles of Pinus sylvestris exposed to ambient air supplemented with NH₃ (53 and 105 µg m⁻³) and fertilized with (NH₄)₂SO₄. Increase with respect to N concentration after exposure to ambient air.

NH ₃ fumiga- tion	(NH ₄) ₂ SO ₄ fertilization	One-year-old needles								Curre year dles	nt- nee-
		Dec 1	988	Feb 1988		May 1988		Aug 1988		Aug 1988	
		FU FE		FU	FE	FU	FE	FU	FE	FU	FE
53	50	99.3	0.7	97.3	2.7	90.0	10.0	91.6	8.4	82.1	17.9
105	50	99.3	0.7	97.9	2.1	82.0	18.0	92.5	7.5	81.2	18.8

tration than one-year-old needles despite the shorter exposure time (Tables 2.3 and 2.4).

The results were the same as in oneyear-old needles, but more signif icant effects of fumigation were o bserved.

Changes in rate of uptake of ferthizer N by needles

The uptake rate of fertilizer N into the needles indicated a pattern of response similar to that of ¹⁵N content (Tables 2.3 and 2.4).

Incorporation of N from fumigation and fertilization

The contribution of NH $_3$ fumigation and $(NH_4)_2SO_4$ fertilization to the N concentration of the needles differed significantly (Table 2.5). In oneyear-old needles, approximately 90% of the increase in N concentration was due to NH_3 fumigation and about 10% to the 50 kg N fertiliz ation. In the current-year needles the percentage contribution from fum igation was 80% compared to 20% from fertilization.

Macronutrient composition

Table 2.6 shows the P , K, Mg and Ca composition of one-year-old and current-year needles. One-year-old needles had lower P, K and Mg and higher Ca concentrations than cu r-rent-year needles. Specifically the K % dry weight was reduced in pr oportion to the NH 4⁺ fertilization.

The concentrations of all m acronutrients in current-year needles were lower in the NH₃ fumigated

Figure 2.3. Changes in water potential (MPa) of needles of Pinus sylvestris exposed to ambient air and fertilized with $(NH_4)_2SO_4$ (kg N ha⁻¹ y⁻¹), on the 17th August 1989. LSD at $\alpha = 0.05$ for all pairwise comparisons.

than in non-fumigated treatments while the total amounts (expressed in mg of dry-weight) were higher. In current-year needles the K % dry weight was not so much affected by NH_4^+ fertilization as in one-year-old needles.

Water potential

The water potential of needles was not affected by NH₃ fumigation (no data given). It decreased signif icantly with N fertilization during the warmest part of the day (Fig 2.3). The daily course of the water pote ntial showed the usual drop during day time from 5:00 to 17:00 h, whereafter recovery occurred from 17:00 to 20:00 h. The drop in p otential was strongly increased by the 100 and 200 kg N application, a lthough full recovery to the control levels occurred between 17:00 and 20:00 h. Table 2.6. Chemical composition of needles of Pinus sylvestris collected after one year exposure to ambient air (AA) and 53 and 105 μ g m⁻³ NH₃ and fertilization with (NH₄)₂SO₄. In the first column the chemical composition is given in mg g⁻¹ DW, in the second column relative to N and in the third column in total amount expressed in mg DW.

NH ₃ fumigat	(./-	SO ₄ Chem iza-	nical co	mposit	ion								
		mg g	¹ DW			% of 1	N (N =	100)		Total amount (mg)			
		Р	K	Mg	Ca	Р	K	Mg	Ca	Р	K	Mg	Ca
One-yea	ar-old needl	es											
AA	0	0.93	2.25	1.40	4.41	9.27	22.43	13.96	43.97	11	25	16	50
AA	50	0.89	1.70	1.71	5.21	8.79	16.73	16.83	51.43	10	18	18	56
AA	100	0.89	1.63	1.55	4.64	8.74	16.04	15.25	45.80	9	17	16	48
AA	200	0.83	1.52	1.42	4.62	6.84	12.49	11.71	38.09	9	16	15	49
53	0	0.82	1.34	1.37	4.14	6.18	10.06	10.32	31.16	6	9	9	28
53	50	0.65	1.29	1.31	3.96	4.62	9.13	9.31	28.11	3	6	6	19
105	0	0.99	1.32	1.28	4.03	7.23	9.64	9.34	29.38	6	8	8	24
105	50	0.76	1.56	1.23	4.18	5.35	11.03	8.72	29.61	7	14	11	36
Current	t-year needl	es											
AA	0	0.89	2.85	1.70	2.70	8.88	28.34	17.00	27.03	18	59	35	56
AA	50	1.25	2.68	2.08	2.84	10.78	23.06	17.93	24.48	23	49	38	52
AA	100	1.37	2.58	2.11	2.61	12.84	24.22	19.85	24.55	27	52	42	52
AA	200	1.13	2.31	1.70	2.20	9.15	18.74	13.82	17.85	32	66	49	63
53	0	0.77	1.98	1.73	2.30	6.70	17.22	15.04	20.00	24	63	55	73
53	50	0.92	1.72	1.75	2.25	7.38	13.83	14.11	18.10	29	54	55	70
105	0	1.06	2.26	1.80	2.26	8.41	17.90	14.25	17.90	35	74	59	74
105	50	0.91	2.14	2.26	2.34	6.65	15.74	17.90	17.17	31	73	56	80

2.5. Discussion

A year of fertilization had hardly any effect on either the needle biomass or the N concentration in the needes. Contribution of the fertilizer N to the total N concentration in the ne edles varied from 2 to 15% depending on the treatment. The small size of this contribution could have several reasons. It might be explained by the distribution of fertilizer N in the tree: in general, roots and trunk have the highest proportion and needles and leaves the lowest (Lalisse & Bardin, 1981; Paavilainen, 1973; Porter et al., 1972). Another reason could be the late moment at which

we applied the fertilizer (August 1988): the N content of the needles can only be increased during the year of the fertilization when fertilizer is applied at the start of the growing season (Paavilainen, 1973). Differences in the shoot-root ratio offer another explanation; Van Dijk et al. (1990) observed that a treatment with 480 kg N ha⁻¹ y⁻¹ applied in rainwater to young coniferous trees increased the shoot-root ratio and decreased the biomass of fine roots and the number of mycorrhizas hence increasing the sensitivity to drought. This hypothesis is su pported by the water po tential results (Fig. 3): the lowest water potential occurred after 100 and 200 kg N fertilization, in the afternoon. This finding agrees with the conclusion of De Kam et al. (1991) who also found a decrease in the water pote ntial with an annual fertilization of 150 and 300 kg N ha $^{-1}$. We confirmed that the lower water potential of the fertilized needles was not due to a higher transpiration resulting from the higher needle biomass. We did not find a decrease in ¹⁵N uptake in fertilized trees (Tables 2.3 and 2.4), indicating that differently from water uptake, the roots still fun ctioned well in N u ptake.

In agreement with Paavilainen (1973) current-year needles had a greater proportion of fertilizer N than one-year-old needles, possibly caused by N translocation from one-year-old to current-year needles and/or a more active transport of ¹⁵N from the roots to current-year ne e-dles. With the emergence of current-

year needles in May the fertilizer N uptake rate by one-year-old needles decreased till the end of the exper iment in August. The stimulation of needle biomass in current-year ne edles (Table 2.2) and the increase in N concentration with time (Fig. 2.2) in trees fumigated with NH₃, confirm that atmospheric NH₃ is used as a foliar supply of N. The NH 3-induced increases in N concentration were not caused by stimulation of the N uptake by the roots (Tables 2.3 and 2.4). Gaseous NH_3 did not increase the uptake of NH_4^+ by the one-yearold needles and it even depressed it in the current-year needles. Needles fumigated with NH₃ had a N conce ntration above 1.3% (Fig. 2.2), a level considered to be sufficient for growth (Van den Burg, 1985); nonfumigated needles were N-deficient (Fig. 2.1). We conclude that uptake of NH₄⁺ is reduced at the root su rface or on the xylem loading when needles get enough N from the gas eous NH₃, which evidently yields a signal to the roots. In the same way, N deficiency in the non-fumigated needles also might provide a signal to the roots to increase the uptake or the transport of N derived from the NH_4^+ fertilizer (see the ¹⁵N concentration and N uptake rate of the ne edles in Tables 2.3 and 2.4). The r esults of Cooper et al. (1986) were similar: when the N supply to the root via the phloem by foliar feeding was increased, the N uptake by the root from the soil was decreased. Therefore, the N status of the shoot may control the N uptake in the roots. Clarkson (1986) suggested

that this shoot-root signal could be a gradient of organic (perhaps specific amino acids) and inorganic co mpounds. Farguhar et al. (1980) and Lemon & Van Houtte (1980) found for ambient concentrations that above the "compensation point" of NH₃ (0.5-5.0 μ g m⁻³), at which net NH₃ uptake occurs, NH₃ was deposited in a linear proportion to ambient NH₃ concentrations, in Zea mays L. and fields of Glycine max L. and Medicago sativa L., respectively. Van Hove et al. (1987) found that NH₃ flux into leaves increases lin early with concentrations up to 400 μ g m⁻³. When Porter et al. (1972) and Whitehead & Lockyer (1987) fumigated corn seedlings and Italian ryegrass, respectively, with NH₃ after isolating the soil-root system from the air in the chamber and fe rtilizing it with NO₃, they found a linear increase in the total plant N as the NH₃ concentration increased. In the present long-term experiment we did not find a linearity between plant N and atmospheric NH₃ in the needles: there was no significant diffe rence between several parameters such as needle biomass (Table 2.2), total N (Fig. 2.2) and ¹⁵N concentrations and fertilizer N uptake rate by needles (Tables 2.3 and 2.4) after fumigation with 53 $\mu g m^{-3} NH_3$ and 105 μ g m⁻³ NH₃. The lack of linea rity could be due to the N concentr ation reached in both fumigations: after 4 months exposure to NH₃, this was close to 2% considered the threshold for damage (Focke, 1991). Hence, in our experiment, 53 μ g m⁻³

NH₃ was already optimal for needle biomass production. The "excess" N provided by 105 µg m⁻³ NH₃ may form part of the cycling-N (Cooper et al., 1986) and/or be stored in other parts of the tree than the ne edles (Bowden et al., 1989; Lalisse & Bardin, 1981). Considering the NH 3 compensation point (max. 5 µg m⁻³ NH_3), we do not expect loss of NH_3 from the needles by volatilization at the concentrations used. In concl usion, linearity between NH₃ flux and NH₃ concentration is limited to short-time studies, since the optimal N concentrations in leaves are soon reached.

Heinsdorf (1991) and Houdijk & Roelofs (1991) found a positive correlation between intensive liv estock breeding, a major source of atmospheric NH₃, and loss of needles. We also found a lower needle biomass in the one-year-old needles of the trees exposed to NH₃, implicating that NH₃ accelerates their fall, thus counteracting the uptake of extra N. The shorter leaf life means an increased production of litter which, because of its high N/C ratio, has a high mineralization rate. Ne edle fall accelerates the cycling of N and the leaching of N to the ground water.

In four months, the N conce ntration in the needles exposed to NH₃ increased by 49% compared with ambient air, while the increase caused by 200 kg N ammonium fertilization was only 8%. This o bservation implies that in our exper iment, NH₃ fertilized faster than $(NH_4)_2SO_4$. Van der Eerden et al. (1992) also found a more marked effect with NH₃ than NH_4^+ on the N concentration in the needles of Pseudotsuga menziesii [Mirb.] Franco, after 13 weeks exposure to 180 µg m^{-3} NH₃ and 500 and 2500 µmol l^{-1} $(NH_4)_2SO_4$. On the one hand, direct foliar uptake of NH₃ can be more efficient for the trees than NH $_4^+$ uptake through the roots, because competition with soil microorga nisms and leaching to the ground water are avoided (Bowden et al., 1989). On the other hand foliar u ptake of NH₃ is disadvantageous for the acid-base regulation of the needle (Raven, 1988).

The general increase in the dry-weight percentage of needles with time was probably caused by an increase in dry matter (as is co mmonly found in Scots pine forests) due to an increase in the starch co ntent of one-year-old needles during spring and early summer (Ericsson, 1979). In August, we found a si gnificantly higher dry-weight percentage in one-year-old needles fum igated with NH₃. This could be explained by a lower water content in the needles since NH₃ increases the shoot-root ratio (Van der Eerden & Pérez-Soba, 1992). Therefore more drought stress in the summer months may be expected.

During the experiment all the treatments showed the same pattern: accumulation of Ca and Mg and a decrease of K and P down to levels of deficiency. The higher contents of P, K, Mg and Ca found in the cu r-rent-year needles exposed to NH₃ could be explained by the higher

needle biomass of the fumigated needles but also by the translocation of nutrients from one-year-old ne edles. Translocation may be caused by faster abscission of dving needles in the fumigated trees and the higher nutrient values of living one-yearold needles compared with those of the current-year needles (Mohren, 1987). NH₃ fumigation significantly decreased the ratios K/N and P/N. These changes in nutrient balance have been related to high N conce ntration in shoots such as those found by Van der Eerden et al. (1992) after fumigating Pseudotsuga menziesii [Mirb.] Franco saplings with 180 µg m^{-3} NH₃.

Toxic effects of NH₃ should always be considered due to the ability of the needles to absorb NH₃ even when the trees are well supplied with N from the soil (Porter et al., 1972; Whitehead & Lockyer, 1987). Toxicity of NH₃ will depend on concentrations of N in the ambient air, plant and soil. The NH₃ concentrations used in our experiment, about 50 and 100 µg m⁻³, are higher than the mean concentrations in the highest NH₃-polluted parts of The Netherlands (25-40 $\mu g m^{-3}$) and comparable to measurements in Germany made by Von Heinsdorf & Krauß (1991), which were 45-65 µg m⁻³. Therefore, an increased emission of NH₃ into the atmosphere, will result in a higher NH₃ uptake by the trees which K/N ratio will decrease, increasing the sensitivity of plants to frost (Dueck et al., 1991) and drought (Van der Eerden et al.,

1991). Hence, plants will be da m-aged more easily.

The present results are relevant for forest management as they ind icate a considerable contribution of atmospheric NH₃ to the N concentration in the needles (Table 2.5) which need to be taken into account in forest fertilization experiments. Extending these findings to crops, atmospheric NH₃ deposition should be estimated in any soil fertilization in order to reduce the N dosage.