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# Evaluating Rice Straw as a Substitute for Barley Straw in Inhibiting Algal Growth in Farm Ponds 

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# Evaluating Rice Straw as a Substitute for Barley Straw in Inhibiting Algal Growth in Farm Ponds 

## Cover Page Footnote

Jacob Maris is a May 2019 honors program graduate with a major in Environmental, Soil, and Water Science. Mary Savin, the faculty co-mentor, is a Professor in the Department of Crop, Soil, and Environmental Sciences.Lisa Wood, the faculty co-mentor, is a Clinical Assistant Professor in the Department of Crop, Soil, and Environmental Sciences.

## Evaluating rice straw as a substitute for barley straw in inhibiting algal growth in farm ponds

## Meet the Student-Author



## Jacob Maris

## Research at a Glance

- Algal blooms can harm aquatic ecosystems and have become more common and severe due to nutrient pollution. Conventional mechanical and chemical methods of algal population control are inefficient and can harm other aquatic organisms.
- Aerobically decomposing barley straw has been shown to inhibit the growth of algal populations. Barley is not a common crop in Arkansas, but other cereal grain straws may release similar chemicals. Rice straw represents a possible eco-friendly, locally sourced form of algal control.
- Based on the results of this study, neither barley straw nor rice straw was effective at algal growth inhibition compared to the control.

I am from Little Rock, Arkansas and graduated from Little Rock Central High School in 2015. In May of 2019, I graduated magna cum laude from the Dale Bumpers College of Agriculture, Food, and Life Science with a degree in Environmental, Soil, and Water Science and a minor in Agricultural Business. Funding was generously provided by the Honors College and Bumpers College to conduct this research and present the results in the undergraduate oral research competition at the ASA, CSSA, SSSA annual conference in Baltimore, Maryland.

I developed a love for both the Razorbacks and the outdoors at a young age. As I learned about our environment and the impact humans have on it in middle and high school, I knew I wanted to make a career out of minimizing that impact. Over the past four years, I have been able to combine these two passions by attending the University of Arkansas. While an undergraduate, I had the opportunity to participate in study abroad programs in Belgium and New Zealand.

Thank you to Dr. Brad Austin for his help in sample analysis and to Dr. Ben Runkle and Dr. Trent Roberts for providing rice straw. I would also like to thank Jody Davis, Brian Austin, Greg Cheshier, Jean Hammack, and LaJoyce Duncan for allowing me to use their ponds in my study.


Jacob analyzes nitrate-N concentrations of pond water samples in the laboratory.

# Evaluating rice straw as a substitute for barley straw in inhibiting algal growth in farm ponds 

Jacob Maris ${ }^{*}$, Mary Savin ${ }^{\dagger}$, and Lisa Wood ${ }^{\S}$


#### Abstract

Algal blooms disrupt aquatic ecosystems and are more common in lakes, ponds, and rivers during the summer months due to nutrient pollution. Livestock production can contribute increased quantities of nutrients to water bodies from runoff of manure. Commonly used mechanical and chemical control methods may have limited success because algae are small and propagate quickly. Barley (Hordeum vulgare) straw has been shown to inhibit the growth of algae as the straw decomposes aerobically in ponds. Therefore, barley represents a natural option for algal biomass control. However, the small amount of barley production in Arkansas limits the availability of barley straw as a solution to control algal blooms locally. Other cereal grain straws may produce similar inhibitory effects during decomposition. Rice (Oryza sativa) is produced in large quantities in Arkansas, making rice straw a locally sourced straw product. The objective of this research was to determine the efficacy of using rice compared to barley straw to inhibit algal growth in freshwater ponds. Data were collected from nine farm ponds, three treated with rice straw, three treated with barley straw, and three without amendment to serve as the experimental control. Dissolved oxygen, pH , nitrate-nitrogen $\left(\mathrm{NO}_{3}^{-}-\mathrm{N}\right)$, dissolved phosphorus ( P ), temperature, and turbidity were measured for 14 weeks from 12 June to 17 September 2018. Algal biomass was measured as chlorophyll-a concentration to evaluate treatment effectiveness over time. Dissolved oxygen was significantly influenced by the main effects of treatment and time. The $\mathrm{NO}_{3}^{-}-\mathrm{N}$ concentration in ponds treated with rice straw was significantly greater than the control and barley treatment. Chlorophyll-a concentrations were variable, and there were no consistent trends through time within a treatment. More research under controlled conditions to understand impacts of abiotic conditions, microbial and algal community compositions, and mode of action of algal inhibition is required before cereal straw can be a reliable, locally sourced method of algal control in farm ponds.


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## Introduction

Algae are present in almost every aquatic ecosystem, play a key ecological role through photosynthesis, and serve as a food source for higher trophic levels. Phytoplankton are free-floating algae that live in the upper layer of stratified ponds and lakes and can grow to large numbers forming algal blooms. These algal blooms, or elevated densities of algal populations, compromise ecosystem health. Increased nutrient concentrations from human activities, such as fertilizer use and livestock production, contribute to more frequent algal blooms (Islami and Filizadeh, 2011). The increase in algal abundance can turn the water color, commonly green in freshwater, and can cause a foul odor. Additionally, dissolved oxygen becomes limited as the algae die and decompose (Kannan and Lenca, 2012). Blue-green algae, while grouped with algae, are photosynthetic bacteria called cyanobacteria (Kannan and Lenca, 2012). Blue-green algae can turn the water green, produce a foul odor, and release cyanotoxins that may be harmful to humans and animals. Attempts to control algae are rarely successful because algae are small and reproduce quickly. Mechanical removal is inefficient and must be repeated periodically, while treatment with chemical algicides can harm non-target organisms (Swistock, 2017).

Barley (Hordeum vulgare) straw can be used as an alternative method of algal control. As barley straw decomposes aerobically, it releases a chemical, or combination of chemicals, that inhibits the growth of green algae and cyanobacteria without harming other aquatic life (Islami and Filizadeh, 2011). While the precise inhibitory chemical is not known, it has been hypothesized that weak peroxides and oxidized polyphenols are responsible for algal growth inhibition (Islami and Filizadeh, 2011). Straw must be placed in ponds 2 to 8 weeks before the algal growing season, depending on water temperature, to give the straw time to begin decomposing (Lembi, 2002). Maximum toxicity to blue-green algae occurs after one month of decomposition and declines over the following months until decomposition is complete (Rice et al., 1980). Decomposition of the barley straw may decrease dissolved oxygen, but the lack of competition for light from algae allows more photosynthesis from higher-order plants (Newman, 2004).

While barley straw has the potential to be an environmentally "clean" form of algal control, there are some concerns regarding the adoption of barley straw for algal control. Barley straw acts as an algistat, rather than an algicide, such that barley straw does not kill existing algal cells but prevents the growth of more algae. Because the Environmental Protection Agency has not certified barley as an algistat, barley straw can be marketed legally only as a home remedy for preventing algal growth (Lembi, 2002). A lo-
gistical challenge to using barley straw in Arkansas is that barley is not a commonly cultivated crop (USDA, 2018). Barley production in the United States is concentrated in the northern midwestern and northwestern states, such as North Dakota, Montana, and Washington, rendering barley straw unavailable to much of the country (Guercio, 2018).

Arkansas is the largest rice (Oryza sativa) producer in the country based on planted area (USDA, 2018). Though studies using cereal straw to prevent algal growth have concentrated on barley straw, other cereal grain straws may be effective substitutes for barley straw because similar chemicals are produced during decomposition (Newman, 2004; Park et al., 2006). The large quantities of rice straw in Arkansas make rice straw favorable when attempting to minimize the cost of algal control. Furthermore, cyanobacterial populations in rice paddies were less dense in the second year of cultivation than in the first year when residues from the first year were left in the paddies, lending support to the hypothesis that rice straw is effective at inhibiting algal growth (Rice et al., 1981).

## Materials and Methods

## Experimental Setup

After presenting the project background, research questions, and anticipated experimental approach to the Ozark Cattlemen's Association and faculty, five volunteers agreed to participate in the project by granting access to their ponds. Nine farm ponds in Washington County were selected for this experiment (Table 1; Fig. 1). For two properties, three ponds were located on one property and each treatment was randomly assigned to a pond on the property. For the last three remaining ponds located on different properties, each treatment was assigned randomly to a pond (Fig. 1). Ponds treated with barley straw were labeled pond B1, B2, and B3. Ponds treated with rice straw were labeled pond R1, R2, and R3, and the three ponds left untreated as an experimental control were labeled pond C 1 , C2, and C3. The surface area was calculated for all ponds by measuring the length and width with a tape measure. Ponds with straw were treated at a rate of $25 \mathrm{~g} / \mathrm{m}^{2}$ with oven-dried straw (Abou El Ella et al., 2007).

The appropriate masses of barley and rice straw were portioned for the respective ponds, cut into pieces approximately 15 cm in length, and placed into plastic $0.5-\mathrm{cm}$ mesh bags. Bags were packed loosely, so water could easily flow through the bag and contact the decomposing straw. When filled with straw, bag volume approximated 90 cm by 55 cm by 40 cm . Pool noodles were tied to the bags with twine to keep the bags afloat in the ponds and promote conditions for aerobic decomposition. Bags were placed on their sides so that the bottom of the bag was at a depth of approximately 20 cm . Each straw bag was anchored to
the pond floor using bricks tied to the ends of string stabilizing placement and evenly spacing bags within ponds. Due to varying pond size, the number of straw bags per pond ranged from one bag in pond R1 to eight bags in pond B2. Brick anchors were attached by a length of string equal to the pond depth at the location of each straw bag. Barley and rice straw bags were placed in ponds on 12 June 2018.

## Sampling

Beginning on 12 June 2018, when the rice and barley straw were placed in each respective pond, water samples were collected weekly from each of the 9 ponds for 14 weeks. Ponds were sampled in the order: R1, B1, C1, R2, B2, R3, B3, C3, C2. Composite samples consisted of 5 individual $125-\mathrm{mL}$ samples collected at a depth of 15 cm ( $625-\mathrm{mL}$ total sample) at regular intervals across a transect dissecting each pond. Individual sample locations corresponded to the following: 1) close to the pond bank, 2) a quarter of the distance across the pond, 3) the center of the pond, 4) three-quarters of the distance across the pond, and 5) at the opposite bank. Samples were collected traversing each pond in an aquatic sampling vessel to prevent water and sediment disturbance. Samples were immediately covered with aluminum foil to prevent further photosynthesis and photodegradation. The final date of sampling was 17 September 2018.

## Chlorophyll-a

To measure chlorophyll-a concentrations, 50 mL from each composite pond sample were filtered in the field using a hand pump and GF/F filter (Whatman, $0.7-\mu \mathrm{m}$ pore size). Filtrate was saved for further filtration for $\mathrm{NO}_{3}^{-}-\mathrm{N}$ and phosphorus analysis. After returning to the laboratory, each filter was soaked in 7 mL of $90 \%$ acetone for 24 hours and stored in a freezer. The extract was analyzed using a Trilogy Laboratory Fluorometer (Turner Designs, San Jose, Calif.). The "Chl-a" module of the fluorometer was calibrated using a stored calibration curve. After samples had been equilibrated to room temperature, extract from each sample ( 3 mL ) was pipetted into a culture tube. Each tube was placed into the fluorometer one at a time. The sample was measured before acidification. After the measurement was complete, 0.1 mL of 0.1 N hydrochloric acid was pipetted into the tube. Following a 90 -second reaction period, the sample was measured after acidification. The acidification step converts all chlorophyll-a to pheophytin, a degradation product of chlorophyll-a, for conversion to a pheophytin-corrected chlorophyll-a concentration measured by the fluorometer. Resulting chlorophyll-a ( $\mu \mathrm{g} / \mathrm{L}$ ) concentrations were recorded.

On week 14 (17 September 2018), nine $50-\mathrm{mL}$ water samples were collected at a depth of 15 cm in pond B 2 to evaluate spatial distribution of chlorophyll-a concentrations. Water samples were collected at distances of 0,3 , and


Fig. 1. Map showing the location of each pond studied in Washington County from Google Earth (2019). Ponds R1, R2, and R3 were treated with rice straw. Ponds B1, B2, and B3 were treated with barley straw. Ponds C1, C2, and C3 were left untreated as a control.

6 m from three separate straw bags for a total of nine samples. Each sample was filtered in the field, prepared, and measured for chlorophyll-a concentration as described previously.

## Nitrate-Nitrogen $\left(\mathrm{NO}_{3}^{-}-\mathrm{N}\right)$ and Dissolved Phosphorus

Using a $0.45-\mu \mathrm{m}$ pore size nylon syringe filter (Cole Parmer, Vernon Hills, Ill.), 10 mL of filtrate from the chlo-rophyll-a procedure were filtered before leaving the field. Two drops of 5-M hydrochloric acid ( HCl ) were added to each sample after filtration to preserve the sample. Nitrate-N concentrations were measured using cadmium reduction and the modified Griess reaction on a Sans-plus segmented-flow autoanalyzer (Skalar Inc, Buford, Ga.) (Baker et al., 2018). The calibration curve was prepared from $0-, 1-, 2-, 5-, 8^{-}$, and $10-\mathrm{mg} / \mathrm{L}$ standards. The filtered and acidified water samples that did not produce an instrument response were recorded as a concentration of 0 $\mathrm{mg} \mathrm{NO}_{3}^{-}-\mathrm{N} / \mathrm{L}$.

Filtered and acidified water samples were sent to the Agriculture Diagnostic Laboratory at the University of Arkansas System Division of Agriculture Agricultural Experiment Station to be analyzed for dissolved phosphorus on a Spectro Arcos inductively coupled plasma-optical emission spectrometer (ICP-OES) (SPECTRO Analytical Instruments, Kleve, Germany).

## Water Quality Parameters

Dissolved oxygen was measured in-situ at each of the 5 locations along the sampling transect, at a depth of 15 cm , using a Lab Quest 2 (Vernier Software \& Technology, Beaverton, Ore.) and Vernier Dissolved Oxygen Probe, and values were averaged for each pond. The pH , temperature, and turbidity were measured on the composite samples in the field using a pH Sensor, Stainless Steel Temperature Probe, and Turbidity Sensor (Vernier Software \& Technology, Beaverton, Ore.).

## Decomposition

To measure the decomposition of straw in each pond, the initial dry weight of straw was measured before adding to each pond. After the 14 weeks, the straw was removed from the ponds, dried in drying ovens at $55^{\circ} \mathrm{C}$ for three weeks, and weighed again. Percent decomposition was calculated by subtracting the final weight from the initial weight, dividing the difference by the initial weight and multiplying by 100 .

## Precipitation

Precipitation data were obtained from the Town Branch at Armstrong Street weather station in Fayetteville, Arkansas on the United States Geological Survey website (USGS, 2018).

## Data Analysis

Chlorophyll-a concentrations were converted to relative percent difference from week 1 concentrations for each pond according to Eq. 1.

$$
\begin{equation*}
\frac{x-x_{0}}{\left(\frac{x+x_{0}}{2}\right)} * 100 \tag{Eq. 1}
\end{equation*}
$$

where $x_{0}$ was the chl-a value in week 1 and $x$ was the chl-a value of the current week.

Averages, standard deviations, and standard error of the mean were calculated each week for the average relative percent difference in chlorophyll-a from week $1, \mathrm{NO}_{3}^{-}$ -N , phosphorus, dissolved oxygen, pH , temperature, and turbidity.

Data organization, graph creation, and data analysis were conducted in Excel 2016 (Microsoft Corp., Redmond, Wash.). Repeated measures analysis of variance (ANOVA) tests were performed on each dependent variable to determine statistical significance at $\alpha=0.05$. Bonferroni posthoc analysis was conducted on variables with significant $P$-values. Statistical analyses were used to determine if dependent variables differed across treatments over time. A $t$-test was used to determine the statistical significance of straw decomposition between rice and barley straw ( $\alpha=$ $0.05)$. A single factor ANOVA was used to determine statistical significance among chlorophyll-a concentrations sampled at increasing distances from the straw bag ( $\alpha=$ 0.05 ). Linear regression was used to determine if dissolved oxygen and temperature changed through time at a $95 \%$ confidence level.

## Results and Discussion

During the 14 weeks, $28.5 \pm 19.3 \%$ (average $\pm$ standard deviation) of the barley straw placed in ponds decomposed, while $43.7 \pm 13.4 \%$ of the rice straw decomposed. Decomposition was not significantly different ( $P=0.26$ ). Barley straw decomposed to the same extent as rice straw $(\sim 40 \%)$ in two of the ponds; however, pond B3 resulted in only a $6.7 \%$ decrease in barley straw. The dissolved oxygen was consistently low, usually between 3 and $4 \mathrm{mg} / \mathrm{L}$ in pond $B 3$. The range for dissolved oxygen concentration in ponds with rice straw and the control was $4.1 \mathrm{mg} / \mathrm{L}$ to $6.5 \mathrm{mg} / \mathrm{L}$ and $3.9 \mathrm{mg} / \mathrm{L}$ to $6.7 \mathrm{mg} / \mathrm{L}$, respectively (Fig. 2). Pond B3 also had a layer of accumulated leaf litter on the bottom of the pond; therefore, aerobic microbial activity in B3, and thus aerobic decomposition of barley straw and production of any allelopathic compounds, may have been more constrained by abiotic conditions in the pond compared to other ponds.

Dissolved oxygen concentration differed among the rice straw and barley straw treatments and the control ( $P<$ 0.001 ). Time affected dissolved oxygen ( $P=0.02$ ). Dis-
solved oxygen concentration in the barley straw treatment ( $4.26 \pm 0.65 \mathrm{mg} / \mathrm{L}$ average $\pm$ standard deviation) was different than dissolved oxygen in both the rice straw treatment ( $5.44 \pm 0.73 \mathrm{mg} / \mathrm{L}$ average $\pm$ standard deviation) and the control $(5.47 \pm 0.76 \mathrm{mg} / \mathrm{L}$ average $\pm$ standard deviation). Dissolved oxygen concentration in week 1 (12 June 2018) differed from weeks $4,5,7$, and 8 . Dissolved oxygen concentration in week 2 (18 June 2018) differed from weeks 4,5 , and 8 , and week 4 differed from weeks 11,12 , and 13 .

Lower dissolved oxygen concentration in the barley straw treatment could decrease the decomposition rate of barley straw in ponds. Although the average dissolved oxygen concentrations varied during the study, the average concentrations remained above $3.6 \mathrm{mg} / \mathrm{L}$, which is sufficient for aerobic decomposition to occur (Cech, 2010). The differences in dissolved oxygen over time could be caused by the changes in water temperatures as the summer progressed. There could have been temporal or spatial locations in at least some of the ponds in which low oxygen concentrations were limiting to the efficacy of straw decomposition to control algal growth.

Initial chlorophyll-a concentrations ranged from 18.8 $\mu \mathrm{g} / \mathrm{L}$ in pond B1 to $457 \mu \mathrm{~g} / \mathrm{L}$ in pond R2. Within the rice treatment, the initial range of concentrations was 436 $\mu \mathrm{g} / \mathrm{L}$. Ponds treated with barley had an initial range of 255 $\mu \mathrm{g} / \mathrm{L}$, and control ponds had an initial concentration range
of $108 \mu \mathrm{~g} / \mathrm{L}$ across the three ponds. The relative percent differences in chlorophyll-a concentrations from week 1 fluctuated through time in all treatments (Fig. 3).

In week 14, the final sample date, both treatments and the control had negative relative percent differences from the week 1 concentration, meaning that there was less algal biomass in week 14 than week 1 in both treatments and the control. In the rice treatment, the relative percent difference ranged from $-90.1 \%$ to $69.8 \%$. In the barley straw treatment, the relative percent differences ranged from $-119 \%$ to $23.7 \%$. In the control group, the relative percent differences raged from $-127 \%$ to $80.2 \%$. The relative percent difference in chlorophyll-a from week 1 of sample collection did not differ statistically between the straw treatments or between the treatments and the control $(P=$ 0.85 , Table 2). During the 14 -week study, there were seven weeks when relative percent differences in all treatments were negative, meaning average chl-a concentrations were less than week 1 , two weeks when the rice straw group was positive, meaning average chl-a showed growth compared to week 1, three weeks when the barley straw group was positive, and four weeks when the control group was positive. Thus, there was no indication of consistent control of algal biomass in either straw treatment, nor was there any consistent trend with algal biomass growth throughout the 14-week experiment ( $P=0.69$, Table 2).


Fig. 2. Dissolved oxygen concentrations (mg/L) in ponds treated with rice straw, barley straw, and no treatment (control) during the fourteen-week study from 12 June 2018 to 17 September 2018. Samples for each treatment were averaged $(n=3)$. Error bars are standard error of the mean.

The variability in chlorophyll-a among ponds within the same treatment could have been caused by environmental factors, such as pond sediment composition, the type and proximity of livestock to the ponds (Table 1), or the flow rate of water within the ponds, factors that were not quantified in this study. For example, pond R2 was spring-fed and feeds into an ephemeral stream. Relative percent difference in chlorophyll-a concentration was negative in all weeks after week 2, indicating algal inhibition throughout the study in pond R2 containing rice straw despite ducks, geese, and cattle having direct access to the pond (Table 1). The movement of water flowing across the pond may have circulated inhibitory chemicals from the decomposing straw throughout the pond. Abou El Ella et al. (2007) controlled algal growth with cereal straw in the Suez Canal, where wave and wind action caused consistent mixing of the water. However, other studies have shown that cereal straw is effective in lentic pond systems as well (Islami and Filizadeh, 2011). Therefore, the efficacy of cereal straw to inhibit algal growth was not expected to be dependent on circulation of water; however, diffusion of inhibitory compounds within farm ponds may be a consideration that requires further investigation.

There was no difference in chlorophyll-a concentrations with distance from the straw bags as measured at 0 , 3 , and 6 m in pond B2 on week 14 (17 September 2018) ( $P$
$=0.49)$. Average chlorophyll-a concentrations were 41.98 $\pm 10.76,41.86 \pm 15.63,52.78 \pm 9.30 \mu \mathrm{~g} / \mathrm{L}$ (average $\pm$ standard deviation) at 0,3 , and 6 m distance from straw bags, respectively. Lack of difference among chlorophyll-a concentrations at different distances from straw bags indicates that diffusion of chemicals dissipating away from the decomposing straw source was not the limiting factor to the efficacy of straw as an algal growth inhibitor in the pond environment.

Water temperature was measured as a factor that could influence decomposition, the algal community, and dissolved oxygen concentrations. Both the treatment ( $P=$ 0.01 , Table 2) and time ( $P<0.001$ ) affected water temperature (Fig. 4). There was no difference among temperatures in the rice and barley treatments; however, both the rice and barley treatments differed from the control. Water temperature in week 2 (18 June 2018) differed significantly from water temperature in week 13 . Water temperature in week 3 (7 July 2018) differed from weeks 7, 10, and 13, week 4 (16 July 2018) differed from weeks 7 and 13, and week 8 (6 August 2018) differed from water temperature in week 13, respectively. Water temperature in week 11 (27 August 2018) differed from weeks $1,5,6,7,9,10$, and 13; week 12 ( 5 September 2018) differed from weeks 7, 9, 10, and 13; week 14 ( 17 September 2018) differed from water temperature in week 7 , respectively.


Fig. 3. Relative percent difference (RPD) from week 1 of chlorophyll-a concentrations ( $\mu \mathrm{g} / \mathrm{L}$ ) in ponds treated with rice straw, barley straw, and no treatment (control) with weekly rainfall data in cm (USGS, 2018) during the 14-week study (12 June 2018 to 17 September 2018). Samples for each treatment were averaged ( $n=3$ ). Error bars depict standard error of the mean.

Table 1. City name, Global Positioning System (GPS) coordinates, surface area, mass of straw added, and the surrounding land use of each pond in the study. Ponds R1, R2, and R3 were treated with rice straw. Ponds B1, B2, and B3 were treated with barley straw. Ponds C1, C2, and C3 were
left untreated as a control.

| Pond ID | City | GPS <br> Coordinates for each Pond | Surface Area | Straw <br> Added | Surrounding Land Use |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\mathbf{m}^{2}$ | kg |  |
| R1 | Farmington, Ark. | $\begin{aligned} & 36^{\circ} 01^{\prime} 49.6^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 14^{\prime} 16.9^{\prime \prime} \mathrm{W} \end{aligned}$ | 230.8 | 5.770 | Horse and donkey pasture with access to pond Occasional cattle |
| R2 | Farmington, Ark. | $\begin{aligned} & 36^{\circ} 03^{\prime} 12.2^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 21^{\prime} 45.2^{\prime \prime} \mathrm{W} \end{aligned}$ | 670.7 | 16.768 | Cattle pasture with access to pond Domestic ducks and geese nesting on pond bank |
| R3 | West Fork, Ark. | $\begin{aligned} & 35^{\circ} 54^{\prime} 42.7^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 07^{\prime} 22.5^{\prime \prime} \mathrm{W} \end{aligned}$ | 1514.9 | 37.870 | Cattle pasture with access to pond Occasional wild ducks in pond |
| B1 | Farmington, Ark. | $\begin{aligned} & 36^{\circ} 01^{\prime} 53.8^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 14^{\prime} 17.5^{\prime \prime} \mathrm{W} \end{aligned}$ | 414.8 | 10.369 | Horse and donkey pasture with access to pond Occasional cattle |
| B2 | Lincoln, Ark. | $\begin{aligned} & 35^{\circ} 56^{\prime} 19.1^{\prime \prime N} \mathrm{~N} \\ & 94^{\circ} 27^{\prime} 08.1^{\prime \prime} \mathrm{W} \end{aligned}$ | 286.5 | 70.911 | Cattle pasture with access to pond |
| B3 | West Fork, Ark. | $\begin{aligned} & 35^{\circ} 54^{\prime} 45.0^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 07^{\prime} 32.2^{\prime \prime} \mathrm{W} \end{aligned}$ | 1631.8 | 40.795 | Cattle pasture with access to pond |
| C1 | Farmington, Ark. | $\begin{aligned} & 36^{\circ} 01^{\prime} 55.4^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 14^{\prime} 11.4^{\prime \prime} \mathrm{W} \end{aligned}$ | 1436.6 | 0 | Horse and donkey pasture with access to pond |
| C2 | Elkins, Ark. | $\begin{aligned} & 36^{\circ} 00^{\prime} 02.3^{\prime \prime N} \\ & 94^{\circ} 00^{\prime} 53.3^{\prime \prime W} \end{aligned}$ | 2251.4 | 0 | Cattle pasture with access to pond Ranging chickens |
| C3 | West Fork, Ark. | $\begin{aligned} & 35^{\circ} 54^{\prime} 22.6^{\prime \prime} \mathrm{N} \\ & 94^{\circ} 07^{\prime} 35.8^{\prime \prime} \mathrm{W} \end{aligned}$ | 2302.7 | 0 | Cattle pasture with access to pond |


| Table 2. Analysis of variance summary of the effects of straw treatment and time |  |  |  |
| :--- | :---: | :---: | :---: |
| on properties measured in pond water. |  |  |  |



Fig. 4. Temperature $\left({ }^{\circ} \mathrm{C}\right)$ in ponds treated with rice straw, barley straw, and no treatment (control) during the 14 -week study from 12 June 2018 to 17 September 2018. Samples for each treatment were averaged ( $n=3$ ).

Error bars are standard error of the mean.


Fig. 5. Nitrate-nitrogen concentrations (mg/L) in ponds treated with rice straw, barley straw, and the control (no treatment) with weekly rainfall data in cm (USGS, 2018) during the 14-week study from 12 June 2018 to 17 September 2018. Samples for each treatment were averaged $(n=3)$. Error bars are standard error of the mean.

The greater temperature in the control ponds could have been due to surrounding land management or an artifact of the sampling procedure. Control ponds lacked tree cover on the banks. More direct sunlight would increase water temperature. The last two ponds sampled each day were both in the control group, so the ponds had more time to warm throughout the day. Randomizing the order in which the ponds were sampled would have controlled for the effect of time of day on water temperatures; however, sampling order was chosen using the most efficient route between ponds to assure all samples could be collected on the same day.

Straw treatment significantly affected $\mathrm{NO}_{3}^{-}-\mathrm{N}(P=$ 0.0003 ) and dissolved phosphorus ( $P=0.02$ ) concentrations, but concentrations did not differ across sampling times ( $P=0.98$ for $\mathrm{NO}_{3}^{-}-\mathrm{N}$ and $P=0.27$ dissolved P measurements across time, respectively, Table 2). Ponds containing rice straw had greater concentrations of $\mathrm{NO}_{3}^{-}-\mathrm{N}$ than ponds containing barley straw or the control (Fig. 5). Average $\mathrm{NO}_{3}^{-}-\mathrm{N}$ concentrations in ponds containing barley straw and the control were 0.013 and $0.009 \mathrm{mg} / \mathrm{L}$, respectively. Average $\mathrm{NO}_{3}^{-}-\mathrm{N}$ concentration in rice strawtreated ponds was $0.599 \mathrm{mg} / \mathrm{L}$. In the rice straw treatment, average dissolved phosphorus concentration was 0.097 $\mathrm{mg} / \mathrm{L}$ and did not differ from the control which averaged
$0.031 \mathrm{mg} / \mathrm{L}$ (Fig. 6). In the barley straw treatment, average phosphorus concentration was $0.123 \mathrm{mg} / \mathrm{L}$, which was greater than the control.

Differences in $\mathrm{NO}_{3}^{-}-\mathrm{N}$ concentrations among treatments could have been due to the type and proximity of livestock to ponds. Cattle had access to all ponds (Table 1). Pond C2 had chickens roaming near the pond; although, the chicken house was downslope from the pond. Horses and donkeys were in fields adjacent to ponds R1, C1, and B1, but were never observed in the water on sampling dates. Pond R2 had domestic ducks and geese that nested on the bank of the pond, and pond R3 occasionally had wild ducks feeding in the pond. The waterfowl in ponds R2 and R 3 might explain the greater concentration of $\mathrm{NO}_{3}^{-}-\mathrm{N}$ in the rice straw treatment. Low $\mathrm{NO}_{3}^{-}-\mathrm{N}$ levels in the control and barley straw ponds could indicate that the conditions necessary for algal growth were not present. The ideal ni-trate-to-phosphate ratio by mass for algal growth is approximately $10: 1$, and concentrations of individual nutrient requirements vary among algal species (Downing and McCauley, 1992). During no week in either treatment or the control was $\mathrm{NO}_{3}^{-}-\mathrm{N}$ concentration great enough to achieve the ideal 10:1 nitrate-to-phosphate ratio for algal growth. Nutrient availability may have contributed to the lack of statistical differences in chlorophyll-a concentrations.


Fig. 6. Dissolved phosphorus concentrations (mg/L) in ponds treated with rice straw, barley straw, and the control (no treatment) with weekly rainfall data in cm (USGS, 2018) during the 14-week study from 12 June 2018 to 17 September 2018. Samples for each treatment were averaged ( $n=3$ ). Error bars are standard error of the mean.

## Conclusions

Due to the lack of differences in chlorophyll-a among the treatments and control, neither rice straw nor barley straw was effective at inhibiting algal growth in the farm ponds studied. Since there was no difference between the rice straw and barley straw treatments, it is unclear if rice straw is as effective as barley straw at inhibiting algal growth. Further research is needed to determine the efficacy of cereal grain straw as a reliable method of algal biomass control.

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