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1 Opportunities and Challenges for an Indonesian

2 Forest Monitoring Network

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130 Key Message:

131	Permanent Sampling Plots (PSPs) are a powerful and reliable methodology to help our							
132	understanding of the diversity and dynamics of tropical forests. Based on the current inventory of							
133	PSPs in Indonesia, there is high potential to establish a long-term collaborative forest monitoring							
134	network. Whilst there are challenges to initiating such a network there are also innumerable benefits							
135	to help us understand and better conserve these exceptionally diverse ecosystems.							
136	Keywords: tropical forests, carbon, data-sharing, dynamics, monitoring							
137	List of abbreviations: NFI = (Indonesian) National Forest Inventory, PSP = permanent sampling							
138	plot, REDD+ = Reducing Emissions from Deforestation and forest Degradation							
139								
140								

141 1. Why monitoring tropical forests is important

142 Tropical forests are arguably the most important terrestrial ecosystems. Whilst occupying 143 around 15 % of the global land area, tropical forests store two-thirds of all the carbon in terrestrial 144 vegetation (Pan et al. 2013) and are the most important above-ground terrestrial carbon sink (Beer et 145 al. 2010; Pan et al. 2011; Soepadmo 1993). They house half the world's biodiversity and provide a 146 wide range of goods, including sources of new medicines, and ecosystem services including clean 147 and sustained water supplies, climate regulation and pollinators for crops (Cámara-Leret et al. 2016; 148 Ghazoul 2015; Peters et al. 1989; Ricketts et al. 2004). If suitably managed, tropical forests can 149 provide economic benefits through ecotourism, non-timber forest products, a sustainable source of 150 timber, and through carbon financing mechanisms for developing tropical countries such as REDD+. 151 Therefore, understanding where, how and why the world's tropical forests are changing is a key 152 question of global importance (Hansen et al. 2013; Pan et al. 2011).

153 The periods over which trees establish, grow and die (tens to hundreds of years) do not make for 154 rapid experimental tests of forest functioning. Instead, direct measurements of stands of trees over 155 long time periods are essential to truly understand forest processes and dynamics (Lutz 2015). 156 Permanent sample plots (PSPs) in which all trees are marked, identified and repeatedly measured 157 provide a series of direct observations on forest condition, dynamics and change over time. As 158 longitudinal data sets, PSPs offer an excellent opportunity to study forest dynamics, and to separate 159 short-term environmental impacts, such as drought, from long-term trends (Condit 1998). A forest 160 monitoring network is a series of PSPs using a consistent protocol - such networks allow an 161 assessment of numerous aspects of forest ecology, including biodiversity, biomass (analogous to 162 carbon stocks), regeneration, dynamics (including succession) and 'health'. Furthermore, forest 163 monitoring networks distributed along large geographical and environmental gradients allow testing 164 for the generality of factors controlling ecosystem functioning with increased statistical power 165 (Craine et al. 2007) and allow space-for-time analyses to project potential impacts of global changes 166 on forests.

Numerous high-impact studies based on PSPs as the fundamental measurement unit have greatly advanced our understanding of the function, biodiversity and evolution of tropical forests. For example, PSPs have provided clear evidence that the tropical forest above-ground carbon stock has been increasing over time (Lewis et al. 2009; Pan et al. 2011; Qie et al. 2017) but that the sink strength into this stock appears to be declining, at least in Amazonia (Breinen et al. 2015). The above studies were conducted in 'undisturbed', i.e. primary, forests but a major proportion of tropical forests have been disturbed by human activities. Fewer PSP networks have been established to study forest recovery from logging (Rutishauser et al. 2015; Sist et al. 2014) or from shifting cultivation (Chazdon
et al. 2016) yet they are also providing valuable data. Furthermore, PSPs contribute vital datasets to
improve our still poor understanding of patterns in tropical tree species richness (Slik et al. 2015; ter
Steege et al. 2013), biogeography (Slik et al. 2018) and evolution (Baker et al. 2014) at multiple scales.
Field data collected on the ground from biogeographically well-replicated PSPs is also a prerequisite
to calibrate remotely-sensed biomass mapping (e.g. Asner et al. 2010; Avitabile et al. 2016; RéjouMéchain et al. 2014).

181 Permanent Sample Plots are a standard method but can be supplemented by biodiversity observing 182 networks such as the transect approach of the Asia-Pacific Biodiversity Observation Network (Yahara 183 et al. 2012, 2014). Larger PSPs (~50 ha), such as those established by the Centre for Tropical Forest 184 Science (CTFS, now ForestGEO), play an important role in furthering our understanding of 185 community ecological patterns as they monitor a larger number of smaller (≥ 1 cm dbh) trees over 186 bigger areas. In contrast, smaller PSPs (usually 1 ha), such as those established by the Amazon 187 Forest Inventory Network (RAINFOR) and the Indonesian National Forest Inventory (see section 2) 188 offer extensive coverage that is more appropriate for a regional-scale forest monitoring network.

189 2. Opportunities from permanent sample plots in Indonesia

190 Indonesia has the third largest area of tropical forest globally (following Brazil and D.R. Congo; 191 FAO 2015) including some of the largest extents of carbon-dense peat swamp forests. However, as 192 with other regions of the world, Indonesia's forests are undergoing rapid change and anthropogenic 193 disturbance (Abood et al. 2014; Gaveau et al. 2014) and around half the country's land area currently 194 supports primary forest (Kementerian Lingkungan Hidup dan Kehutanan 2015b; Margono et al. 195 2014). The forests of western Indonesia are highly productive and the dominant trees, the 196 dipterocarps (Brearley et al. 2016), have been favoured as commercial timber trees for many years 197 leading to the majority of accessible forests being brought into timber production. By contrast, the 198 forests of eastern Indonesia (especially Papua) contain few dipterocarps and remain more intact 199 owing to the rugged topography and isolation. More recent challenges include droughts and fires 200 associated with El Niño that have had marked impacts upon forest functioning (Page & Hooijer 2016; 201 Slik 2004) and increasing forest fragmentation (Qie et al. 2017), yet large-scale analyses that test for 202 such impacts across Indonesian forests are largely absent.

203 Numerous PSPs have been established across Indonesia over the last c. 60 years but not all have been 204 maintained continuously. The earliest PSPs were established during the late Dutch colonial era, but 205 they were mostly in plantation forests to study tree growth and timber yield (Hart 1928; Von Wulfing 206 1938). Among the first PSPs established in primary forest was the 1-ha plot set-up by Willem Meijer 207 (1959) to study the ecology of Gunung Gede's montane forests. Since then, PSPs have played an 208 important role in silvicultural research such as the STREK (Silvicultural Techniques for the 209 Regeneration of Logged-over Forest in East Kalimantan) project (Bertault & Kadir 1998). The 210 Indonesian National Forest Inventory (NFI) is a national program initiated by the Indonesia Ministry 211 of Forestry in 1989 (and implemented by the Directorate General of Forestry Planning) utilizing PSPs. 212 Through this program, PSPs were established systematically with a 20 x 20 km grid across forested 213 areas in Indonesia (< 1000 m above sea level) with the primary objective to monitor the growth of 214 timber stocks. In total, 2735 1-ha PSPs were established, although not all have been monitored on 215 more than one occasion (Kementerian Kehutanan 1996). Depending on the location, the NFI plots 216 were not necessarily located in logging concessions but all logging companies were required to 217 establish PSPs for monitoring growth and yield. In addition to monitoring timber growth and yield, 218 data from these PSPs has provided a basis for estimating carbon stocks and changes associated with 219 land-use change and forest management activities (Kementerian Lingkungan Hidup dan Kehutanan 220 2016; Krisnawati et al. 2014, 2015).

221 Despite the large-scale coverage of Indonesia's NFI, the limited scientific access NFI offers to its data 222 and the few large-scale analyses that have resulted from the NFI's dataset limit our understanding of 223 the composition and functioning of Indonesia's tropical forests. Given the current threats to 224 Indonesia's forests, it is important that Indonesian and foreign scientists collaborate, with a 225 consolidated scientist-led forest monitoring network having the flexibility to address ecological 226 questions in a democratised and collaborative fashion, to jointly establish PSPs and analyse large 227 datasets spanning Indonesia's forests. To date, at least 150 ha of PSPs (besides those in the NFI) have 228 been established in primary forest, and are still maintained, in Indonesia (Table 1; Figures 1a & 2). 229 Although these PSPs have different sizes, re-measurement intervals and measurement protocols 230 making direct comparisons challenging, they offer a starting point for developing an Indonesian 231 forest monitoring network with a standardised protocol. The density of sampling across the whole 232 of Indonesia is only about 3.4 ha of plots per 10⁶ ha of primary forest and there are clear differences 233 in sampling density between different geographical regions (Table 1). The highest density (ratio of 234 plot area to primary forest area) of PSPs, by an order of magnitude, is found in Java and Bali (Table 235 1). Although the total area of PSPs is modest, the area of primary forest remaining is particularly 236 low on these islands leading to an overall very high sampling density. Of the outer islands, 237 Kalimantan has a high density of sampling – likely due to this being the centre of production forest 238 logging activity coupled with interest in its exceptional biodiversity since the times of early colonial 239 explorers. Sumatra has a similar sampling density and has also been heavily exploited for timber in 240 the past. Maluku also has a high sampling density but this is largely confined to Seram only. 241 Sulawesi and Nusa Tenggara have sampling densities comparable to the mean for the whole of 242 Indonesia (although note that there are only 2.5 ha of plots in Nusa Tenggara). Sampling density 243 for Papua is, by far, the lowest among the Indonesian islands; this is partly due to the large remaining 244 area of forest combined with difficulties in establishing PSPs in areas with challenging access. Of 245 these PSPs, nearly half have been measured on more than one occasion, thereby markedly increasing 246 their value for assessing forest functioning, with the median monitoring period for those measured 247 more than once being 8 years and the longest being 50 years (Fig. 2b). About half of the plots that 248 have been measured on more than one occasion are in Kalimantan (e.g. Qie et al. 2017) so the total 249 monitoring effort (plot area x monitoring length) at around 1300 ha years is an order of magnitude 250 greater than Java + Bali, Maluku, Sulawesi or Sumatra; none of the PSPs in Nusa Tenggara or Papua 251 have been re-measured (Fig. 2c). In addition, there are over 100 ha of PSPs in disturbed forest (Fig. 252 1b); many of these are forests that have been logged; in this case, the geographical foci are Kalimantan 253 and Sumatra that have historically been important for timber and, secondarily, in Papua where 254 logging activities are currently expanding.

From the brief analysis above, it is clear that key geographical gaps exist mainly in eastern Indonesia particularly for Maluku (excepting Seram), Nusa Tenggara and Papua. In terms of climate, many areas of drier forest are under-represented (e.g. Timor), as is montane forest and forest over edaphic variants (such as *kerangas* or ultramafic geology). There are some PSPs found in peat swamp forests but many have been burnt or otherwise disturbed in recent years.

260 3. Challenges facing an Indonesian forest monitoring network

261 3.1 Methods

Our aim here is not to provide a protocol or critique of methods for PSPs as this has been done
in previous work (Alder & Synott 1992; Burslem & Ledo 2015; Condit 1998; Ledo 2015; Phillips et al.
2016; Sheil 1995) but to note concerns with particular relevance to the Indonesian situation.

265

Plot size: Too many PSPs reported in the Indonesian literature are simply too small to provide a generalisation of the area they study. Small plots (e.g. 0.04 ha) might be useful when installed in a series (e.g. 25) to provide data on forest biodiversity that does not require accurate scaling-up to larger areas. However, for a more in-depth assessment of forest biodiversity, the larger the area sampled, the greater the number of species captured due to a large number of rare species (Plotkin et al. 2000). Of the PSPs noted in our analysis, the median size is 0.25 ha whilst the most frequently sized plot is 1 ha (Figure 2a), which is comparable to forest monitoring networks on other continents (Brienen et 9

273 al. 2015; Lewis et al. 2009; Phillips et al. 2009, 2016). Small plots cannot accurately predict forest 274 biomass when scaled-up to a larger area due to a high edge:interior ratio that elevates the relative 275 importance of marginal boundary decisions (Burslem & Ledo 2015), a high coefficient of variation 276 between plots, and the likelihood they will not represent all forest stages (e.g. gap, building and 277 mature, sensu Whitmore 1998). Calibration of remote sensing data for large-scale forest biomass 278 mapping is more accurate if the PSPs can be ground-truthed accurately, which also requires larger 279 plots (Avitabile et al. 2016; Réjou-Méchain et al. 2015). Finally, small plots are also prone to the 280 'majestic effect' where researchers may unconsciously select pristine forest with 'majestic' large trees 281 and avoid disturbed areas (Sheil 1995).

282 Frequency of measurement: Whilst the definition of a PSP is that trees will be re-measured at some point 283 in time, re-measurement intervals are not always regular. A typical re-measurement interval is five 284 years as this allows increases in tree size to be seen more easily. Whilst intervals of four to ten years 285 are appropriate for most recording purposes of PSPs (Sheil 1995), an increasing census period leads 286 to a greater likelihood of unobserved growth and therefore an underestimation of forest productivity 287 (Talbot et al. 2014). In cases of annual censuses, this will allow much better predictions of forest 288 dynamics in relation to annual climate fluctuations (Clark et al. 2010). Dendrometer bands are a 289 possible inexpensive alternative to increase measurement frequency (Anemaet & Middleton 2013), 290 but require much greater time investment at installation; such bands can also avoid errors due to 291 changes of the point of measurement. Of course, regularity of re-measurement depends upon plot 292 security and accessibility, and funding is a key determinant of frequency of fieldwork activities (see 293 section 3.3).

294 Parameters measured: Trunk diameter at breast height (usually 1.3 m) is the key parameter measured 295 as this can be incorporated into allometric equations to estimate tree and stand biomass (Chave et al. 296 2014); including tree height and crown size has been shown to increase accuracy of such equations 297 (Goodman et al. 2014). This is especially needed for dipterocarps that show different architectural 298 patterns compared to other tropical trees (i.e. taller for a given diameter: Banin et al. 2012). Forests 299 in Indonesia cover not only a wide range of soil and climatic types both within and across islands, 300 but also represent a great biogeographical range. Due to variable architectures that require local 301 height-diameter models for accurate biomass calculation, tree height data collected within plots are 302 extremely useful to improve biomass estimates (Ledo et al. 2016; Sullivan et al. 2018).

303 3.2 *Taxonomy*

For assessment of species distributions and monitoring, accurate taxonomy, comparable among
 plots, is paramount. Good taxonomy is clearly challenging as PSPs often contain a large proportion

306 of sterile individuals. Indonesia is fortunate in having a large and well-maintained national 307 herbarium (Herbarium Bogoriense; BO) and a number of regional herbaria but many PSP 308 investigators do not routinely collect voucher specimens but rely on vernacular names instead. 309 Taxonomy takes on extra importance in a forest monitoring network where the aim is to make 310 comparisons among plots, but technological advances have a key role to play here (Baker et al. 2017; 311 Webb et al. 2010). While some Indonesian tree genera are reasonably well known, for example the 312 commercially important dipterocarps (Ashton 2004) many large genera such as Syzygium (Myrtaceae) 313 and Diospyros (Ebenaceae) have not been monographed. Similarly, digitization of herbarium sheets 314 at BO is ongoing but progress remains slow.

315 Vouchers for morphotypes can be made available across sites permitting analysis of distribution of 316 taxa without any formal species names, but obtaining the species name increases the value of the 317 voucher. Challenges for the taxonomy of PSP trees must be taken seriously, and we recommend the 318 following: i) make physical voucher collections of several specimens for each morphotype especially 319 where variation appears to be high and collect silica gel-dried samples for subsequent DNA 320 barcoding; ii) carry out routine visits to PSPs to collect fertile specimens as they become available; iii) 321 take high-quality photographs of the fresh vouchers (Webb et al. 2010) and share images and 322 metadata online; iv) cross-match vouchers and images across different sites to both validate formal 323 species name and provide distribution information; v) avoid the use of vernacular names, except as 324 an early step in the determination process yet value the experience of parataxonomists in the field 325 and technicians in herbaria; and vi) publish details of how taxon names were acquired, and give a 326 level of confidence in each formal name. Overall, it is far more useful to publish voucher collection 327 codes, images, morphotype codes and matches of morphotypes to images at other sites than to simply 328 list a botanical name with no additional information. Detailed primary data will also greatly assist 329 taxonomic specialists in the future as they work on the large, complex genera of Indonesian trees.

330 3.3 Funding

331 Funding presents a perennial challenge for forest ecological work, particularly in developing 332 countries. Within Indonesia, PSP censuses are not considered as applied research, which receive 333 priority for funding, although NFI plots have been allocated governmental funding. Current 334 funding opportunities through the development of the Indonesian Science Fund (DIPI) and via the 335 UK Newton Fund are positive in this regard. There is also the potential for knowledge-exchange 336 partnerships with logging companies who may fund PSPs in their concessions although, as funders, 337 they may consider themselves data owners (see section 3.4). REDD+ programmes bring similar 338 opportunities for knowledge exchange and funding (Gibbs et al. 2007). Longer-term collaborations 339 between Indonesian researchers, companies and NGOs coupled with leading international expertise

340 are needed. Importantly, PSPs need to be locally owned, and international funding should be

341 invested for pump-priming and capacity-building in order to stimulate long-term funding input from

342 Indonesian sources into tropical forest monitoring.

343 3.4 Data-sharing

344 Developing an integrated picture on changes in forest functioning and biodiversity across a 345 forest monitoring network requires the willingness to share data among researchers. Nevertheless, 346 data-sharing can present various challenges. There are a number of data-sharing models in tropical 347 ecology, ranging from the informal to the formal with rigid data-sharing arrangements such as 348 ForestPlots (López-González et al. 2011). What is shared can vary from whole plot data to only the 349 numbers required for a particular analysis. Issues over intellectual property are of considerable 350 concern and unwillingness to share data is often linked to concerns about the loss of control over 351 such data and the lack of professional recognition or reward (Enke et al. 2012; Fecher et al. 2015). 352 Furthermore, clarifying who is the 'owner' of data is essential. In some cases, the funder (often a 353 logging company) may claim ownership, in others, such as the Indonesian NFI, public access to the 354 data is limited. Any forest monitoring network needs clear guidelines on the sharing, use and 355 publication of shared data and an obvious reward system for sharing (i.e. co-authorship).

356 Although in-country data owners will regularly be included as co-authors in large-scale data 357 analyses, the lead authors have almost always been researchers from extra-tropical countries. 358 Echoing the sentiments of Ruslandi et al. (2014), we note that simply 'out-sourcing' data analysis to 359 extra-tropical researchers is still far from the goal of building local research capacity. Lack of 360 institutional support and incentive may deter tropical scientists from becoming leading authors, but 361 this appears to be changing lately with Indonesian institutions increasingly rewarding staff 362 publishing in international journals. Investing in capacity-building and knowledge exchange to 363 support Indonesian scientists to take leadership roles in agenda setting is also important in the 364 medium term.

365 3.5 Land tenure and community engagement

366 Once a series of PSPs has been established it is important to maintain a commitment to re-measure 367 plots and obtain funding to do so. However, the location and accessibility of plots needs to be 368 considered for long-term measurements. Ideally, plot locations should not be too remote to make 369 accessibility challenging and not too close to settlements put plots at risk from disturbances. If new 370 PSPs are installed, there should be secure land tenure (Soraya 2011) to offer protection from land-use 371 change and fire risk - particularly in peat swamp forests (Page & Hooijer 2016). Of the PSPs noted 372 (Table 1; Figures 1 & 2), less than half are within formally protected areas (e.g. National Parks or 373 Nature Reserves); of those that are not, the presence of researchers may help in protecting them to 374 some degree (Laurance 2013). In areas where forest land-use classifications may jeopardise studies, 375 it may be possible to re-designate land classifications (e.g. Kawasan Hutan Dengan Tujuan Khusus 376 or 'Special Use Forests'). Local stakeholder engagement is key, and local communities should be 377 considered as valuable collaborators who value the presence of PSPs and can be employed to collect 378 good quality data (Theilade et al. 2015). There are multiple opportunities for synergies between 379 local communities, logging companies and scientists, with NGOs often in a strong position to act as 380 facilitators. Still, unless direct payments to forest owners are established for missed opportunities 381 of economic development, communities may well continue to prefer the economic benefits offered 382 by logging companies over those from researchers or conservationists (Novotny 2010).

383 4. Translating results from PSPs to forest policy and conservation

Quantification and assessment of carbon stocks in forests underpins international policies to mitigate carbon dioxide emissions such as the REDD+ program (Gibbs et al. 2007) and the recommendations of the Intergovernmental Panel on Climate Change (Watson et al. 2000). For example, Indonesia's forest reference emission level submitted to United Nations Framework Convention on Climate Change (Kementerian Lingkungan Hidup dan Kehutanan 2015a, refined in 2016) utilized NFI data as the primary source to generate information on carbon stocks (and thus emissions from forest change).

391 It is essential to understand not only carbon stocks in tropical forests through time but also the 392 response of tropical forest to climate change and develop policies accordingly. Information from 393 PSPs will allow us to determine whether Indonesian forests are sinks or sources of carbon and have 394 the potential to help us understand the factors driving carbon stock changes. To derive national 395 policies, information from PSPs needs to be combined with data on land use and land-use change, 396 which is accessible through remote sensing data or national inventories.

In addition, tropical forests are also key repositories of global biodiversity, genetic resources and important ecosystem services for local communities. Reducing biodiversity loss is a target of the United Nations Convention on Biological Diversity (Pereira et al. 2013) which is not only relevant from an aesthetic point of view, but can also threaten ecosystem functioning (Duffy 2009). Permanent sample plot data will foster a better understanding of the autecology, distribution and rarity of tree species and they also have the potential to obtain measures of biodiversity of various

- 403 taxonomic groups at multiple scales and to link the abundances of each of these with one another.
- 404 All of the above are needed to enhance Indonesia's conservation planning efforts and manage forests
- 405 in a way that allows biodiversity to flourish in this exceptionally biodiverse country.
- 406
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- 410

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- 587 **Table 1.** Areas of forested land and sampled by permanent sample plots (PSPs) in primary forest
- 588 (excluding the National Forest Inventory) on major islands of Indonesia. Data on land and
- 589 forest area taken from Kementerian Lingkungan Hidup dan Kehutanan (2015b).

Island(s)	Land area (10º ha)	Total forested area (10 ⁶ ha)	Primary forest area (10° ha)	Total PSP area (ha)	PSP/forest area ratio**
Java (+ Bali)	13.95	3.37	0.08	9.0	113.0
Sumatra	47.16	14.07	4.49	38.0	8.5
Kalimantan	52.96	27.58	9.80	82.1	8.4
Sulawesi	18.53	9.47	3.91	12.3	3.1
Nusa Tenggara*	6.76	2.84	0.68	2.5	3.7
Maluku	7.77	5.11	0.96	12.3	12.8
Papua	40.79	34.06	26.15	2.0	0.1
Total	187.92	96.50	46.07	158.1	3.4

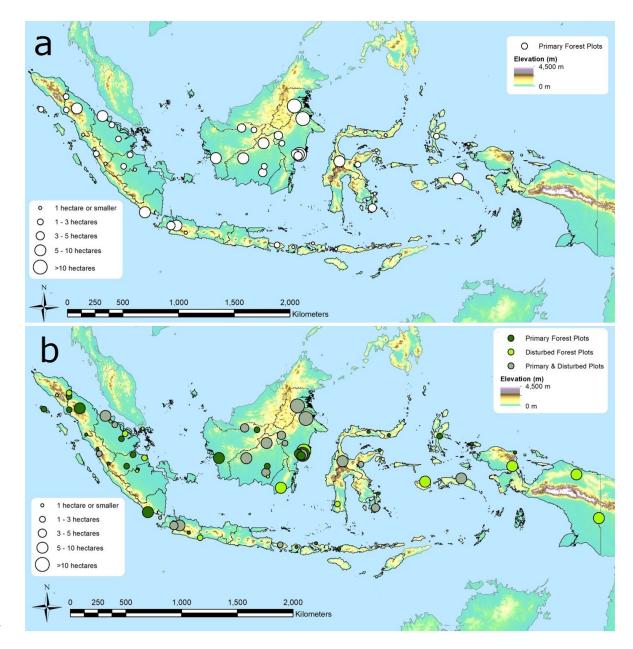
590 * Excluding Bali, which is included with Java due to their biogeographical affinity.

591 ** Area of permanent sampling plots (ha) per 10⁶ ha of primary forest.

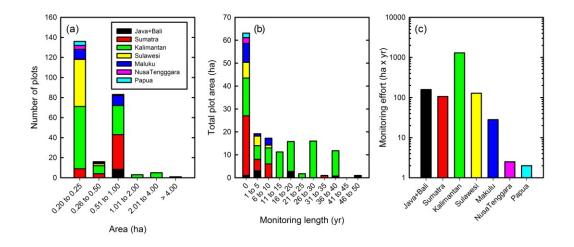
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- 595 Figure 1. (a) Locations of primary forest and (b) primary and disturbed permanent sampling
- 596 plots (PSPs) in Indonesia (excluding the National Forest Inventory).



- 599 Figure 2. (a) Plot areas, (b) total plot area under different lengths of monitoring and (c) total
- 600 monitoring effort (i.e. sum of area multiplied by monitoring length for each plot) for permanent
- 601 sample plots (PSPs) in primary forest (excluding the National Forest Inventory) on major islands
- 602 of Indonesia. Note that plots only measured once are given a monitoring length of one year
- 603 and also note the logarithmic scale for panel (c).



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