1	Running title: Mutual benefits along plants life-span
2	Title:
3	Benefits for nurse and facilitated plants emerge when interactions are considered
4	along the entire life-span
5	
6	
7	Alicia Montesinos-Navarro <sup>1</sup> , Isabelle Storer <sup>2</sup> and Rocío Perez-Barrales <sup>2</sup>
8	
9	1 Centro de Investigaciones sobre Desertificación (CIDE, CSIC-UV-GV), Carretera
10	de Moncada-Náquera Km 4.5 46113 Moncada, Valencia, Spain.
11	Alicia Montesinos-Navarro (ali.montesinos@gmail.com)
12	
13	2 School of Biological Sciences, King Henry Building, King Henry I Street,
14	University of Portsmouth, Portsmouth PO1 2DY, UK.
15	Rocío Pérez-Barrales (rocio.barrales@port.ac.uk);
16	Isabelle Storer ( <u>isabelle.storer@myport.ac.uk</u> )
17	
18	Correspondence author:
19	Alicia Montesinos-Navarro (ali.montesinos@gmail.com)
20	ORCID-ID: 0000-0003-4656-0321
21	Centro de Investigaciones sobre Desertificación (CIDE, CSIC-UV-GV), Carretera de
22	Moncada-Náquera Km 4.5 46113 Moncada, Valencia, Spain.
23	

#### 24 ABSTRACT

25 The structure of plant communities is often influenced by facilitative interactions 26 where 'facilitated' plants gain a benefit from growing associated with 'nurse' plants. 27 Facilitation has been mostly studied from the facilitated plant's perspective, and 28 bidirectional effects between nurse and facilitated plants have received less attention. 29 We hypothesized that reciprocal benefits in plant-plant interactions may emerge when 30 interactions are considered along the life-span of the plants involved. Over one spring, 31 we selected five species with similar life-form and growth strategy, and using a full 32 factorial design, we compared different fitness components along the plants' life-span 33 (seedling establishment, juvenile growth and reproductive investment in adult plants). 34 We compared: a) plants growing in solitary stands and associated with other plants in 35 vegetation patches; and b) plants that originally functioned as nurse plant (the largest 36 plant of the vegetation patch) and as facilitated (not the largest plant of the vegetation 37 patch). Plants growing in vegetation patches displayed higher seedling establishment 38 and juvenile growth compared to solitary conspecific plants. At a later developmental 39 stage, nurse plants in vegetation patches experienced higher reproductive investment 40 (measured as flower production relative to plant size) compared to solitary plants. In 41 contrast, the originally facilitated plants showed similar reproductive investment 42 compared to their solitary pair of similar size. Facilitation might be a complex 43 interaction in which reciprocal benefits for both facilitated and nurse plants can be 44 detected when interactions are considered along the plants' life-span. Our results 45 suggest that mutual benefits in plant-plant interactions could be important to sustain 46 diversity in plant communities, but they appeared overlooked and deserve further 47 attention.

48 Keywords: Facilitative interactions, gypsum, long-term interactions, nurse plants'

49 benefits, plant community, priority effects

- 50
- 51

# 52 INTRODUCTION

53 Community assembly processes strongly influence the relative abundance of species 54 in communities and the spatial distribution of biodiversity (HilleRisLambers et al. 55 2012, Mittelbach and Schemske 2015). Plant facilitative interactions occur if at least 56 one of the interacting species gains some benefit (facilitated) from growing in 57 association with other (nurse) species, and no harm is caused to either (Callaway 58 2007). During the period of establishment, plants commonly benefit by growing 59 associated with a nurse plant, but this positive interaction can shift to competitive 60 once the facilitated plant reaches maturity (Miriti 2006, Schiffers and Tielb rger 2006, Armas and Pugnaire 2009). A less explored component of facilitation is 61 62 whether changes in the shared environment caused by adult facilitated plants benefit 63 the original nurse plant (i.e. the plant that arrived first to the shared environment). In 64 stressful environments with limited resources, this could result in positive feedbacks 65 between plants, although these might be detectable at different developmental or 66 demographic stages for the nurse and the facilitated plants. Unfortunately, the 67 dynamics of positive feedbacks along the entire life-span and different demographic 68 stages of plants remains largely unexplored.

69

Plant-plant facilitative interactions are commonly assessed from the perspective of
benefits gained by the facilitated species (Callaway 2007). The analysis of costs and
benefits for both facilitated and nurse plants allowed identification of shifts from

73 facilitation to competition associated with plant growth (Valiente-Banuet et al. 1991, 74 Armas and Pugnaire 2009). However, recent studies demonstrated that the benefits of 75 plant-plant interactions can be mutual (Sortibrán et al. 2014, Tirado et al. 2015), 76 which could be explained if the interacting plants shift their role, e.g. if the 77 beneficiary species becomes benefactor as it grows and becomes an adult plant. For 78 example, during an early life stage, seedling establishment and growth of juvenile 79 plants increases if plants grow under the stress-ameliorated micro-environment 80 provided by the nurse plant (Pugnaire et al. 1996, Gómez-Aparicio et al. 2004, 81 Callaway 2007, Castillo et al. 2010); as plants become adults, they modify the micro-82 environment (Navarro-Cano et al. 2015), which in turns benefits those plants that 83 share it. Adult plants can provide fungal mutualisms, increase nutrient turn over by 84 altering soil microbiota, and accumulate nutrients, all of which benefit other plants 85 (Montesinos-Navarro et al. 2012, 2016, Rodríguez-Echeverría et al. 2013, Lozano et 86 al. 2014, Navarro-Cano et al. 2014, Sortibrán et al. 2014), including the original nurse 87 plant. Therefore, as adult plant, the originally facilitated plant can enrich the micro-88 environment shared with the nurse plant. The nurse plant might overall gain larger net 89 benefits compared to the adult facilitated plant, since larger plants are more efficient 90 at capturing resources (Goldberg and Landa 1991, Keddy et al. 2002, Vogt et al. 91 2009). Nurse and adult facilitated plants might compete for resources during specific 92 periods, but in the long term, the association between plants is likely to remain if the 93 net effects are positive for both nurse and facilitated plant, and overcome the potential 94 costs (Valiente-Banuet and Verdú 2007). Shifting roles from beneficiary to benefactor 95 and vice versa might strengthen species coexistence, ultimately increasing 96 biodiversity at the level of local communities.

98 In this study, we focus on a Mediterranean plant community growing on 99 resource-poor gypsum soils to investigate whether mutual benefits for the nurse 100 and the facilitated plants may emerge throughout the lifespan of plants. We 101 compare the performance measures between isolated and aggregated plant 102 species at different developmental stages (i.e., juveniles and adult plants) and 103 expect that both the facilitated and the original nurse plant species can benefit 104 from the interaction, the former at an early developmental stage and the later as 105 an adult (Fig. 1). Overall, our approach contributes to a better understanding of 106 the temporal dynamics of plant facilitative interactions, and provides insights into mutual benefits between nurse and facilitated plants, otherwise overlooked 107 108 in the plant-plant facilitation literature.

109

#### 110 MATERIALS AND METHODS

111

112 Study site

113 The study was conducted in spring-summer 2016 in a semi-arid plant community in 114 the SE of Spain, 2.21 km NE of Petrer in the province of Alicante (38°29'52.36"N, 115 0°44'37.62"W), in an area of 4708 m<sup>2</sup> with a 30% of vegetation cover (Delalandre and 116 Montesinos-Navarro 2018). The average annual temperature and precipitation of the 117 region is 22.5°C and 381 ml respectively (Cueva 1994). The study site is an outcrop of gypsum soil surrounded by clay and loam. Gypsum soils are physically unstable 118 119 due to lack of plasticity, cohesion, and aggregation (Romão and Escudero 2005), 120 properties that prevent root penetration of the surface crust after seed germination and 121 limit establishment of plant juveniles (Meyer 1986). Gypsum soils generally lack 122 important plant nutrients such as phosphorus, nitrate and potassium (Meyer 1986).

123 Altogether these physical and chemical properties make gypsum soils stressful124 environments for plants.

125 The most common plant species in the study site were chamaephytes, including 126 Teucrium libanitis Schreber, Thymus vulgaris, Th. moroderi, Fumana ericoides, 127 Helianthemum squamatum (Jacq.) Dum. Cours., H. syriacum L., Helichrysum 128 stoechas, Herniaria fruticosa, and Sedum sediforme. Shrubs such as Cistus clusii, 129 Ononis tridentata and Rosmarinus officinalis were scattered and uncommon. We 130 focused the study on the most abundant species, namely T. libanitis, Th. moroderi, H. 131 squamatum, H. syriacum and He. Stoechas, because they are all dwarf shrubs with 132 similar size and similar traits related to growth strategies (trait values compiled from 133 BROT data base, presented in Table 1; Paula et al. 2009, Tavcsanouglu and Pausas 134 2018). Therefore, differences among these species are likely to emerge in relation to 135 temporal differences in plant establishment, especially in gypsum soils, where 136 secondary growth shows large inter-site and small interspecific differences (Olano et 137 al. 2011).

138

139 Sampling design

140 Over one spring, we studied five species with similar life-form so that their size could 141 be used as proxy of the sequence order of plant establishment in vegetation patches. 142 We compared performance traits relevant to different life stages between plants that 143 appeared growing in association with other species (i.e. in vegetation patches where 144 canopies overlapped) and solitary, in order to estimate the net benefits (or costs) of 145 growing in vegetation patches. For seedlings and juvenile plants, we quantified the 146 frequency of establishment and growth of juveniles respectively. In adult plants, we 147 determined whether net flower production (relative to plant height) varied with the 148 role that the species played in the interaction (nurse or facilitated) (Fig. 1). Therefore, 149 we quantified the net benefit in pairs of species where the associated individual was 150 the largest in the vegetation patch (assumed to be the original nurse plant), and in 151 pairs where it was not the largest (although frequency of establishment and juvenile 152 growth might have been facilitated by the nurse plant at an earlier life stage). We 153 focused on key demographic parameters that influence population dynamics, namely 154 survival, growth and reproduction (Caswell 2001), and selected plant performance 155 components that inform about these parameters. We used plant height as proxy of 156 either growth and/or survival, and flower production as a proxy of reproduction. 157 Then, we followed a widely used approach to quantify the intensity of plant-plant 158 interactions, which consists of comparing plant performance between treatments 159 differing in the presence (e.g. associated) or absence (e.g. solitary) of neighbors 160 (Goldberg et al. 1999, Holzapfel and Mahall 1999, Pugnaire and Luque 2001, 161 Cavieres et al. 2008, Sch□b et al. 2014b, García-Cervigón et al. 2016, Pueyo et al. 162 2016, Llambí et al. 2017). We applied this general approach for the following fitness 163 components along the plants' life-span.

164

## 165 Frequency of juvenile recruitment

Recruitment and spatial distribution of juvenile plants (e.g. early life development stage) was studied by sampling 111 plots of 1.5 m ×1.5 m randomly distributed across the study site. We quantified the number of juvenile plants growing in vegetation patches with other plants, and growing solitarily. We considered juvenile plants those with soft stems and without signs of stem lignification, and vegetation patches as groups of plants with overlapping canopies. We calculated the percentage of the total area covered by plants by measuring the maximum and minimum diameter of the patch per plot, and approximated the patch area to an ellipse. We used binomial tests to compare the observed number of juveniles in the two conditions (i.e. growing in association with other plants in patches and solitarily), with those expected based on the proportion of area covered by plants, using the "binom.test" function implemented in R version 3.2.3 (R Core Team 2015). The p-values were corrected for multiple testing through the false discovery rate (FDR) method, using the "p.adjust" function.

179

180 Juvenile growth

In an area of 30 m<sup>2</sup>, we compared height of juvenile plants growing in patches and 181 182 isolated to test if performance increased in vegetation patches. Plant height was used 183 as proxy of performance, driven either by an increase in growth rates and/or improved 184 survival. We used a general linear mixed model and included height as dependent 185 variable, growing condition (e.g. in association with other plants and in solitary 186 stands) as fixed effect, and species identity as random effect. The analysis was 187 conducted using the "lme4" package (Bates et al. 2015), the significance of the fixed 188 effect was assessed using "ImerTest" package (Kuznetsova et al. 2017), and the mean 189 estimate of the levels was retrieved using "effects" package (Fox and others 2003), all 190 implemented in R version 3.2.1 (R Core Team 2015).

191

192 Flower production of adult plants

We measured plant height and flower production in 200 adult individuals. We selected plants distributed in a paired design (hereafter set) so that two plants of the same species appeared growing in each condition, i.e. in vegetation patches and in solitary stands within 1 m distance from each other. In half of the sets, the plant growing in association was the largest of the vegetation patch (assumed to be 198 originally the nurse plant), whereas in the other half the adult plant was not the largest 199 (assumed to be the originally facilitated plant as seedling) (Fig. 1). For T. libanitis and 200 Th. moroderi we counted the number of inflorescence rather than number of flowers 201 because flowers of these species are too small to obtain accurate counts in the field. 202 We targeted five species and selected 20 sets per species (40 plants per species), 203 accumulating a total of 100 sets (sets included paired-individuals, one from the 204 vegetation patch and one from solitary stands, of a similar size; 200 plants in total). In 205 some cases, sets of different species shared the same vegetation patch, and this was 206 incorporated in the analyses (see below).

207 Because plant size and/or age (both likely correlated with height) affects flower 208 production, we conducted a linear regression for each species to predict flower 209 production based on plant height. We extracted the residuals and used them as 210 measure of corrected flower production to remove plant size effects on the response 211 variable. Within each set, we estimated the difference between the corrected flower 212 production of plants growing in association and its solitary paired-plant. We used this 213 measure as proxy of benefits gained (if positive) and costs (if negative) of growing 214 associated with other plants in a vegetation patch. Two conspecific plants within each 215 set could slightly differ in height, but using the residuals of flower production 216 regressed against plant height ensured that detectable differences between plants in 217 association and its solitary paired-plant were independent of height. We used a 218 general linear mixed model to test if differences in corrected flower production could 219 be explained by the fixed factor "role". "Role" indicated whether the associated plant 220 was assigned as originally nurse or facilitated plant (largest and not the largest plant 221 of the patch respectively). In the analysis, plant species and vegetation patch were 222 included as random factors. All statistical analyses were performed using the package "Ime4" (Bates et al. 2015), the significance of the fixed effects was assessed using
"ImerTest" package (Kuznetsova et al. 2017), and the mean estimates and standard
error of the two levels of the fixed factor (largest and non-largest plant) were obtained
using "effects" package (Fox and others 2003), both implemented in R version 3.3.1
(R Core Team 2015).

228

229 RESULTS

230

231 Juvenile recruitment

In the 250 m<sup>2</sup> area sampled, plants covered only 15% of the area surveyed, but the estimated probability of finding juveniles in vegetation patches was threefold of solitary juveniles for all plant species (0.52-0.59, Table 2). The binomial test showed that the proportion of juveniles growing associated with vegetation patches was significantly larger than expected by chance (Table 2).

237

238 Juvenile growth

Height was recorded in 314 juvenile plants in 30 m<sup>2</sup> (154 of *H. syriacum*, 82 of *H. squamatum*, 28 of *Th. moroderi*, 21 of *T. libanitis* and 19 of *He. stoechas*). The height of juvenile plants was significantly explained by growing condition, whether the plants appeared growing associated *vs.* in solitary stands (F-value =  $6.07_{1, 309}$ , p-value = 0.01), with an average height (mean  $\pm$  SE) of  $2.88 \pm 0.42$  cm in vegetation patches and  $2.23 \pm 0.40$  in solitary stands (Fig. 2). These differences represent an increase of 29% in height for juvenile plants in vegetation patches.

246

247 Flower production

We detected differences in corrected flower production in relation to the role assigned 248 249 to plants. Nurse plants (i.e. the largest individuals in vegetation patches) showed 250 significantly higher benefits than facilitated conspecifics (not the largest) (N=100, p-251 value= 0.003; Fig. 3). Specifically, differences in corrected flower production 252 between the largest plant in the patch and its paired-solitary conspecific was positive 253 and statistically significant (72.38  $\pm$  25.22; p-value = 0.02). In contrast, when the 254 associated plant was not the largest of the patch, flower production was similar 255 between the plant in the vegetation patch and the paired-solitary conspecific (-11.02  $\pm$ 256 25.22; p-value = 0.67).

257

258 DISCUSSION

259 From its foundation, coexistence theory emphasized competitive interactions to 260 predict species composition in communities (Chesson 2000). More recently, facilitation has been incorporated to this framework (Bulleri et al. 2016) under the 261 262 view of being an unidirectional process with benefits for only one plant involved in 263 the association (Callaway 2007). The integration of a bidirectional and temporal 264 perspective to plant-plant interactions has provided evidence of shifts from facilitation 265 to competition, as facilitated plants became adults (Armas and Pugnaire 2009). 266 However, it remains untested whether adult facilitated plants could change their role 267 and become benefactors, providing benefits to other plants, including the original 268 Our study suggests that facilitation is maybe a complex process nurse plant. 269 maintained throughout the life cycle of the plants involved in the interaction. Here we 270 propose that as facilitated plants grow and become adults, both plants involved in the 271 interaction could benefit from being associated with each other. It is therefore important to investigate the potential benefits gained at different developmental stages 272

273 by identifying meaningful performance measures relevant to those life stages (e.g. 274 survival and growth in juvenile vs. flower production as proxy of reproductive 275 investment in adults). Previous research on plant-plant facilitation quantified changes 276 in the magnitude and sign of interactions along the ontogeny of facilitated plant 277 (Armas and Pugnaire 2005) and the effect of neighbours on its population dynamics 278 (Miriti et al. 2001, Butterfield et al. 2010). However, to our knowledge, the proposed 279 conceptual frame of mutual benefits for facilitated and nurse species along their life 280 span has not been integrated in previous studies, and therefore the potential (delayed) 281 benefits for nurse plants may have been overseen.

282 We investigated plant-plant interactions by testing whether plants growing in 283 vegetation patches gained benefits compared to those growing in solitary stands. Our 284 analysis revealed that more juveniles recruited associated to other plants in vegetation 285 patches compared to solitary, while at a later developmental stage, the original nurse 286 plant showed an increased flower production compared to solitary conspecifics of 287 similar size. In similar systems, vegetation patches provide a microenvironment richer 288 in nutrients and with greater diversity of microbiota compared to bare ground 289 (Navarro-Cano et al. 2014). Further research is required to assess whether an 290 improvement of soil properties (e.g. moisture, fertility, enriched microbiota, among 291 others) could be the underlying mechanism of facilitation.

From a theoretical perspective, while experimental manipulations (i.e. plant removal, addition of nutrients) seem appropriate to evaluate mechanisms for plant facilitation, it might not be feasible to conduct them in unique edaphic islands with endemic plant species (Escudero et al. 2015). In addition, manipulations do not necessarily guarantee the removal of all the effects induced by the association. Hence, nonexperimental approaches like ours might be appropriate to study plant-plant 298 interactions. Nevertheless, in non-experimental approaches other biotic or abiotic 299 factors could have influenced the performance measurements. Micro-environmental 300 features of the landscape might affect soil nutrient content or water availability, 301 conditioning the growth and reproduction of plants stablishing in different microsites 302 (Lundhold 2009, Stover and Henry 2018). The relatively small area (0.5 Ha) and low 303 slope of the study site selected should have minimized this potential micro-304 environmental variation. In addition, the paired experimental design to compare 305 flower production between associated and solitary plants, where each pair is no more 306 than one meter apart, should also account for potential micro-environmental variation 307 present within our study site. In the case of early-stages performance measurements 308 other factors such as a seed-trap effect might have influenced recruitment of juveniles 309 in vegetation patches. However, it is unlikely that juvenile growth in vegetation 310 patches could be attributed to seed-trap effects. Hence, other benefits, such as higher 311 moisture or nutrient availability influenced by the presence of other plants, might be 312 present in these environments. We identified some large individuals growing solitary, 313 maybe reflecting a limited seed dispersal and colonization of all suitable micro-314 environments, as well as variability in the plant-plant interactions outcomes as a 315 balance between competition and facilitation.

316

Facilitative priority effects occur when the establishment of a species modifies the biotic or abiotic environment in ways that favour the establishment of subsequently arriving species (Fukami et al. 2005). The time of arrival of different species can in turn affect performance of later arriving species, which is often reflected in the community structure (von Gillhaussen et al. 2014). Although the stage-dependent responses to plant-plant interactions studied are not equivalent to documenting 323 temporal dynamics, our results describe a pattern that might support the presence of 324 different effects of facilitative interactions based on the arrival time of the species 325 involved. In our study, this is supported by the fact that the same species showed 326 different responses to growing associated vs. solitary, depending on whether they 327 were the largest plant in the vegetation patch. Although larger size relative to other 328 plant species in the vegetation patch may not necessarily imply an earlier arrival time, 329 this might be likely considering that the plant species selected share similar traits 330 related to growth strategies.

331 Plants that establish first can influence recruitment, growth or reproductive success of 332 subsequently arriving species (Fukami 2015), and the mechanisms commonly 333 invoked include niche pre-emption (e.g. a reduction of available niche for plants with 334 similar requirements), or niche modification (i.e. soil legacies) (Weiner 1990, Grman 335 and Suding 2010, Cuddington 2011, van der Putten et al. 2013, Helsen et al. 2016). 336 Different mechanisms of soil modification can give rise to beneficial effects, for 337 example, by increasing nutrient availability or soil mutualisms (Burkle and Belote 338 2015), but this modification might depend on the size of the established plants. Our 339 study suggests that adult plants, particularly when they are the largest of the 340 vegetation patch, could benefit from the association with other plants. We measured 341 this by comparing flower production between vegetation patches and solitary 342 condition. Flower production and associated reproductive traits are often sensitive to 343 nutrient availability particularly in stressful environments (Muñoz et al. 2005, Burkle 344 and Irwin 2009). It is possible that plants established in a vegetation patch modify soil 345 properties improving the microenvironment and resource availability of the patch 346 (Navarro-Cano et al. 2014). While this can benefit all plants in the patch, the original 347 nurse plant (i.e. the largest in the patch), which probably presents a more developed root system than other adult plants, may be more efficient at acquiring resources, and
thus more likely to gain greater benefits than relatively smaller plants (Violle et al.
2009, Wang et al. 2010). Nevertheless, further research will help improving a
mechanistic understanding of these interesting patterns.

352

353 Reciprocal benefits in plant facilitative interactions have received little attention, but 354 the interest is increasingly growing (Pugnaire et al. 1996, Holzapfel and Mahall 1999, 355 Sortibrán et al. 2014, Tirado et al. 2015). Plants interact on a multidimensional scale, 356 exerting positive and negative effects on their partners depending on the fitness 357 component considered (i.e. germination, seed production, flower density, fruit set) 358 (Maestre et al. 2003, Sch b et al. 2014a). However, plant traits vary across the life-359 span, implying that the benefits of harbouring complementary suites of traits might 360 only be detectable when facilitative interactions are considered along the entire life-361 span of plants (Montesinos-Navarro et al. 2019). At earlier developmental stages, 362 seedlings or juvenile plants might be unable to substantially modify the soil 363 conditions in which they grow (Sortibrán et al. 2014), but their capacity increases 364 along the ontogeny of the plant (Navarro-Cano et al. 2015). In contrast, adult plants 365 can reduce insolation, enhance nutrient cycling through litter leaching, alter microbial 366 communities, and enhance the performance and reproductive success of neighbour 367 plants (Vetaas 1992, Pugnaire et al. 1996, Montesinos-Navarro et al. 2012, Navarro-368 Cano et al. 2014, Rodríguez-Echeverría et al. 2016). Future research should 369 incorporate a wider and functional perspective of other reproductive estimates, such 370 as fruit and seed production and quality, or probability of pollen delivery and arrival, 371 all of which positively correlate with larger floral displays (Karron and Mitchell 2011). Measuring performance at different life stages allows identifying benefits that 372

373 might be delayed and emerge only when plant-plant interactions are considered along374 the life-span of the species involved.

375

376 In conclusion, this study provides a novel view of the stage-dependent benefits that both nurse and facilitated plants can gain from facilitative interactions. Mutual 377 378 benefits in facilitative interactions might promote long-term co-existence in plant communities, particularly in stressful environments, and create positive feedbacks in 379 plant-plant interactions. Nevertheless, more research is required to provide a 380 381 functional perspective and discover the underlying mechanisms that create positive 382 feedbacks between plants, and how these structure plant diversity at the community 383 level.

384

# 385 ACKNOWLEDEMENTS

AMN was supported by a postdoctoral contract from the Spanish Ministry of 386 387 Economy and Competitiveness (FPDI-2013-16266; IJCI-2015-23498). Financial 388 support was also provided by the regional government Generalitat Valenciana 389 (GV/2016/187). RPB was supported by a Percy Sladen Memorial Grant from The 390 Linnean Society. IS was supported by Erasmus + with a work placement fellowship to 391 conduct her undergraduate dissertation project. We are especially grateful to Maya 392 Zomer, who kindly read the manuscript and provided insightful comments, and to M. 393 Verdú and P. García-Fayos, for the support and hosting provided to IS during the 394 development of her project.

395

### **396** LITERATURE CITED

397 398	Armas, C., and F. I. Pugnaire. 2005. Plant interactions govern population dynamics in a semi-arid plant community. Journal of Ecology 93:978–989.
399 400 401 402 403	Armas, C., and F. I. Pugnaire. 2009. Ontogenetic shifts in interactions of two dominant shrub species in a semi-arid coastal sand dune system. Journal of Vegetation Science 20:535–546.
404 405 406	Bates, D., M. Mächler, B. Bolker, and S. Walker. 2015. Fitting linear mixed-effects models using lme4. Journal of Statistical Software 67:1–48.
407 408 409 410	Bulleri, F., J. F. Bruno, B. R. Silliman, and J. J. Stachowicz. 2016. Facilitation and the niche: implications for coexistence, range shifts and ecosystem functioning. Functional Ecology 30:70–78.
411 412 413	Burkle, L. A., and R. T. Belote. 2015. Soil mutualists modify priority effects on plant productivity, diversity, and composition. Applied vegetation science 18:332–342.
414 415 416 417	Burkle, L. A., and R.E. Irwin. 2009. The effects of nutrient addition on floral characters and pollination in two subalpine plants, Ipomopsis aggregata and Linum lewisii. Plant Ecology 203: 83-98.
417 418 419 420 421	Butterfield, B. J., J. L. Betancourt, R. M. Turner, and J. M. Briggs. 2010. Facilitation drives 65 years of vegetation change in the Sonoran Desert. Ecology 91:1132–1139.
421 422 423 424	Callaway, R. M. 2007. Positive interactions and interdependence in plant communities. Springer.
425 426 427	Castillo, J. P., M. Verdú, and A. Valiente-Banuet. 2010. Neighborhood phylodiversity affects plant performance. Ecology 91:3656–3663.
428 429 430	Caswell, H. 2001. Matrix Population Models: Construction, Analysis, and interpretation. 2nd ed., Sinauer Associates, Sunderland, Masssachusetts.
431 432 433 434 425	Cavieres, L., C. Quiroz, and M. Molina-Montenegro. 2008. Facilitation of the non- native Taraxacum officinale by native nurse cushion species in the high Andes of central Chile: are there differences between nurses? Functional Ecology 22:148– 156.
435 436 437 438	Chesson, P. 2000. Mechanisms of maintenance of species diversity. Annual review of Ecology and Systematics 31:343–366.
439 440 441	Cuddington, K. 2011. Legacy effects: the persistent impact of ecological interactions. Biol. Theory 6:203–210.
442 443 444	Cueva, A. J. P. 1994. Atlas climático de la Comunidad Valenciana (1961-1990). Consellería de vivienda, obras públicas y vertebración del territorio.
445 446	Delalandre, L., and A. Montesinos-Navarro. 2018. Can co-occurrence networks predict plant-plant interactions in a semi-arid gypsum community? Perspectives in

447	plant ecology, evolution and systematics 31:36–43.
448	Equidance A. S. Delacio, F. T. Magatan and A. I. Lymping 2015. Plant life on
449 450	gynsum: a review of its multiple facets. Biological Reviews 90:1–18
451	gypsum, a review of its multiple facets. Diological Reviews 90.1 10.
452 453	Fox, J., and others. 2003. Effect displays in R for generalised linear models. Journal of statistical software 8:1–27
454	of statistical software 0.1 27.
455	Fukami, T. 2015. Historical contingency in community assembly: integrating niches,
456 457 458	species pools, and priority effects. Annual Review of Ecology, Evolution, and Systematics 46:1–23.
450 159	Fukami T. T. Martiin Bezemer S. R. Mortimer and W. H. Putten 2005 Species
460 461	divergence and trait convergence in experimental plant community assembly. Ecology Letters 8:1283–1290.
462	
463 464 465	García-Cervigón, A. I., J. M. Iriondo, J. C. Linares, and J. M. Olano. 2016. Disentangling facilitation along the life cycle: Impacts of plant-plant interactions at vegetative and reproductive stages in a Mediterranean forb. Front. Plant Sci. 7:129.
466	
467 468 469	Von Gillhaussen, P., U. Rascher, N. D. Jablonowski, C. Plückers, C. Beierkuhnlein, and V. M. Temperton. 2014. Priority effects of time of arrival of plant functional groups override sowing interval or density effects: a grassland experiment. PLoS
4/0	One 9:e86906.
4/1	Coldborg D. E. and V. Landa 1001 Compatitive offect and regnance: hierarchies
472 473 474	and correlated traits in the early stages of competition. The Journal of Ecology 79.1013–1030
475	
476	Goldberg, D. E., T. Rajaniemi, J. Gurevitch, and A. StewartOaten. 1999. Empirical
477 478	approaches to quantifying interaction intensity: competition and facilitation along productivity gradients. Ecology 80:1118–1131.
479	
480 481 482	Gómez-Aparicio, L., R. Zamora, J. M. Gómez, J. A. Hódar, J. Castro, and E. Baraza. 2004. Applying plant facilitation to forest restoration: a meta-analysis of the use of shrubs as nurse plants. Ecological Applications 14:1128–1138
483	sinuos as narse plants. Deological ripplications 11.1120 1150.
484	Grman, E., and K. N. Suding, 2010, Within-Year Soil Legacies Contribute to Strong
485	Priority Effects of Exotics on Native California Grassland Communities.
486	Restoration Ecology.
487	
488	Helsen, K., M. Hermy, and O. Honnay. 2016. A test of priority effect persistence in
489	semi-natural grasslands through the removal of plant functional groups during
490	community assembly. BMC ecology 16:22.
491	
492	HilleKisLambers, J., P. Adler, W. Harpole, J. Levine, and M. Mayfield. 2012.
493 101	Retrinking community assembly through the lens of coexistence theory. Annual Powiow of Foology, Evolution, and Systematics 42:227
494 495	Keview of Ecology, Evolution, and Systematics 45.227.
175	

496 497	Holzapfel, C., and B. E. Mahall. 1999. Bidirectional facilitation and interference between shrubs and annuals in the Mojave Desert. Ecology 80:1747–1761.
498	
499 500	Karron, J. D., and R. J. Mitchell. 2011. Effects of floral display size on male and female reproductive success in Mimulus ringens. Annals of botany 109:563–570.
501 502 503 504 505	Keddy, P., K. Nielsen, E. Weiher, and R. Lawson. 2002. Relative competitive performance of 63 species of terrestrial herbaceous plants. Journal of Vegetation Science 13:5–16.
505 506 507 508	Kuznetsova, A., P. B. Brockhoff, and R. H. Christensen. 2017. ImerTest Package: Tests in Linear Mixed Effects Models. Journal of Statistical Software 82:1–26.
500 509 510 511 512	Llambí, L. D., N. Hupp, A. Saez, and R. Callaway. 2017. Reciprocal interactions between a facilitator, natives, and exotics in tropical alpine plant communities. Perspectives in Plant Ecology, Evolution and Systematics.
512 513 514 515 516	Lozano, Y. M., S. Hortal, C. Armas, and F. I. Pugnaire. 2014. Interactions among soil, plants, and microorganisms drive secondary succession in a dry environment. Soil Biology and Biochemistry.
510 517 518 519	Lundholm, J. T. 2009. Plant species diversity and environmental heterogeneity: spatial scale and competing hypotheses. Journal of Vegetation Science 20: 377- 391
520 521 522 522	Maestre, F. T., S. Bautista, and J. Cortina. 2003. Positive, negative, and net effects in grass-shrub interactions in Mediterranean semiarid grasslands. Ecology 84:3186–3197.
523 524 525	Meyer, S. E. 1986. The ecology of gypsophile endemism in the eastern Mojave Desert. Ecology 67:1303–1313.
526 527 528	Miriti, M. N. 2006. Ontogenetic shift from facilitation to competition in a desert shrub. Journal of Ecology 94:973–979.
529 530 531 532 533	Miriti, M. N., S. Joseph Wright, and H. F. Howe. 2001. The effects of neighbors on the demography of a dominant desert shrub (Ambrosia dumosa). Ecological Monographs 71:491–509.
534 535 536	Mittelbach, G. G., and D. W. Schemske. 2015. Ecological and evolutionary perspectives on community assembly. Trends in Ecology & Evolution.
537 538 539 540	Montesinos-Navarro, A., J. G. Segarra-Moragues, A. Valiente-Banuet, and M. Verdú. 2012. Plant facilitation occurs between species differing in their associated arbuscular mycorrhizal fungi. New Phytologist 196:835–844.
541 542 543 544 545	Montesinos-Navarro, A., M. Verdú, J. I. Querejeta, L. Sortibrán, and A. Valiente- Banuet. 2016. Soil fungi promote nitrogen transfer among plants involved in long- lasting facilitative interactions. Perspectives in Plant Ecology, Evolution and Systematics 18:45–51.

546 547 548 549	Montesinos-Navarro, A., M. Verdú, J.I. Querejeta, A. Valiente-Banuet. 2019. Nurse shrubs can receive water stored in the parenchyma of their facilitated columnar cacti. Journal of Arid Environments 165:10-15.
550 551 552 553	Munoz, A. A., C. Celedon-Neghme, L.A. Cavieres, and M.T. Arroyo. 2005. Bottom- up effects of nutrient availability on flower production, pollinator visitation, and seed output in a high-Andean shrub. Oecologia 143: 126-135.
554 555 556 557	Navarro-Cano, J. A., M. Goberna, A. Valiente-Banuet, A. Montesinos-Navarro, C. García, and M. Verdú. 2014. Plant phylodiversity enhances soil microbial productivity in facilitation-driven communities. Oecologia 174:909–920.
558 559 560	Navarro-Cano, J. A., M. Verdú, C. García, and M. Goberna. 2015. What nurse shrubs can do for barren soils: rapid productivity shifts associated with a 40 years ontogenetic gradient. Plant Soil 388:197–209.
562 563 564 565	Olano, J. M., M. Eugenio, and A. Escudero. 2011. Site effect is stronger than species identity in driving demographic responses of Helianthemum (Cistaceae) shrubs in gypsum environments. American journal of botany 98:1016–1023.
565 566 567 568 569	Paula, S., M. Arianoutsou, D. Kazanis, Ç. Tavsanoglu, F. Lloret, C. Buhk, F. Ojeda, B. Luna, J. Moreno, A. Rodrigo, and others. 2009. Fire-related traits for plant species of the Mediterranean Basin. Ecology 90:1420–1420.
570 571 572 573	Pueyo, Y., D. Moret-Fernández, A. I. Arroyo, A. Frutos, S. Kéfi, H. Saiz, R. Charte, M. de la L. Giner, and C. L. Alados. 2016. Plant nurse effects rely on combined hydrological and ecological components in a semiarid ecosystem. Ecosphere 7.
574 575 576 577	Pugnaire, F., P. Haase, J. Puigdefábregas, M. Cueto, S. Clark, and L. Incoll. 1996. Facilitation and succession under the canopy of a leguminous shrub, Retama sphaerocarpa, in a semi-arid environment in south-east Spain. Oikos 76:455–464.
578 579	Pugnaire, F. I., and M. T. Luque. 2001. Changes in plant interactions along a gradient of environmental stress. Oikos 93:42–49
581 582 583 584 585	Van der Putten, W. H., R. D. Bardgett, J. D. Bever, T. M. Bezemer, B. B. Casper, T. Fukami, P. Kardol, J. N. Klironomos, A. Kulmatiski, J. A. Schweitzer, and others. 2013. Plant-soil feedbacks: the past, the present and future challenges. Journal of Ecology 101:265–276.
586 587 588	R Core Team. 2015. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. 2013.
589 590 591 592	Rodríguez-Echeverría, S., C. Armas, N. Pistón, S. Hortal, and F. I. Pugnaire. 2013. A role for below-ground biota in plant-plant facilitation. Journal of Ecology 101:1420–1428.
593 594 595	Rodríguez-Echeverría, S., Y. M. Lozano, and R. D. Bardgett. 2016. Influence of soil microbiota in nurse plant systems. Functional Ecology 30:30–40.

596 597	Romão, R. L., and A. Escudero. 2005. Gypsum physical soil crusts and the existence of gypsophytes in semi-arid central Spain. Plant Ecology 181:127–137.
598 599 600	Schiffers, K., and K. Tielb rger. 2006. Ontogenetic shifts in interactions among annual plants. Journal of Ecology 94:336–341.
601 602 603 604 605 606 607	Sch□b, C., R. Michalet, L. A. Cavieres, F. I. Pugnaire, R. W. Brooker, B. J. Butterfield, B. J. Cook, Z. Kikvidze, C. J. Lortie, S. Xiao, and others. 2014a. A global analysis of bidirectional interactions in alpine plant communities shows facilitators experiencing strong reciprocal fitness costs. New Phytologist 202:95– 105.
608 609 610 611	Sch□b, C., I. Prieto, C. Armas, and F. I. Pugnaire. 2014b. Consequences of facilitation: one plant's benefit is another plant's cost. Functional Ecology 28:500– 508.
612 613 614 615	Sortibrán, L., M. Verdú, and A. Valiente-Banuet. 2014. Nurses experience reciprocal fitness benefits from their distantly related facilitated plants. Perspectives in Plant Ecology, Evolution and Systematics 16:228–235.
616 617 618 610	Stover, H. J., and H. A. Henry, 2018. Soil homogenization and microedges: perspectives on soil - based drivers of plant diversity and ecosystem processes. Ecosphere 9(6): e02289.
620 621	Tavcsanouglu, Ç., and J. G. Pausas. 2018. A functional trait database for Mediterranean Basin plants. Scientific data 5:180135.
623 624 625	Tirado, R., K. A. Bråthen, and F. I. Pugnaire. 2015. Mutual positive effects between shrubs in an arid ecosystem. Sci. Rep. 5.
626 627 628	Valiente-Banuet, A., and M. Verdú. 2007. Facilitation can increase the phylogenetic diversity of plant communities. Ecology Letters 10:1029–36.
629 630 631 632	Valiente-Banuet, A., F. Vite, and J. A. Zavala-Hurtado. 1991. Interaction between the cactus Neobuxbaumia tetetzo and the nurse shrub Mimosa luisana. Journal of Vegetation Science 2:11–14.
633 634 635	Vetaas, O. R. 1992. Micro-site effects of trees and shrubs in dry savannas. Journal of vegetation science 3:337–344.
636 637 638 639	Violle, C., E. Garnier, J. Lecoeur, C. Roumet, C. Podeur, A. Blanchard, and ML. Navas. 2009. Competition, traits and resource depletion in plant communities. Oecologia 160:747–755.
640 641 642 643	Vogt, D. R., D. J. Murrell, and P. Stoll. 2009. Testing spatial theories of plant coexistence: no consistent differences in intra-and interspecific interaction distances. The American Naturalist 175:73–84.
644 645	Wang, P., T. Stieglitz, D. W. Zhou, and J. F. Cahill Jr. 2010. Are competitive effect and response two sides of the same coin, or fundamentally different? Functional

- Ecology 24:196–207.

- Weiner, J. 1990. Asymmetric competition in plant populations. Trends in ecology & evolution 5:360–364.

Table 1. Trait values related to plant growth for the five focal species of our study

654 compiled from BROT data base (Paula et al. 2009, Tavesanouglu and Pausas 2018):

655 Helianthemum squamatum (H. squamatum), Helianthemum syriacum (H. syriacum),

656 Helichrysum stoechas (He. stoechas), Teucrium libanitis (T. libanitis), Thymus

657 moroderi (Th. moroderi). The similarity in trait values (number of individuals (N),

658 mean and standard deviation (sd)) across the plant species selected in the study,

supports that they have similar growth strategies. All data comes from specimens

660 collected in Eastern Spain.

Species traits	Ν	mean	sd		
Height (m)					
H. squamatum	2	0.26	0.20		
H. syriacum	5	0.46	0.24		
He. stoechas	4	0.48	0.14		
T. libanitis	1	0.12	-		
Th. moroderi	1	0.10	-		
Leaf area (mm <sup>2</sup> )					
H. squamatum	1	69.54	-		
H. syriacum	2	44.80	14.42		
He. stoechas	3	39.66	32.73		
T. libanitis	1	23.44	-		
Th. moroderi	1	4.50	-		
Leaf Nitrogen con	tent (	(mg/g)			
H. squamatum	1	13.39	-		
H. syriacum	1	12.74	-		
He. stoechas	3	12.82	0.79		
Root depth (m)					
H. squamatum	2	0.44	0.30		
H. syriacum	2	0.39	0.39		
He. stoechas	2	0.33	0.27		
T. libanitis	1	0.17	-		
Th. moroderi	1	0.13	-		
Specific leaf area (mm <sup>2</sup> /mg)					
H. squamatum	1	4.56	-		
H. syriacum	3	4.84	1.47		
He. stoechas	3	10.15	2.04		
T. libanitis	1	3.41			
Th. moroderi	1	6.30	-		

663 Table 2. Binomial tests assessing differences in the establishment of juveniles in 664 vegetation patches and bare ground for the five focal species. The expected proportion 665 of juveniles growing associated with other plants was 0.15, based on the percentage of cover of plants in the 30m<sup>2</sup> plot area sampled. For each species, the total number of 666 667 juveniles observed (Total N. juveniles), the number of juveniles growing associated 668 with other plants (N. juv. associated), the estimated probability of growing associated 669 based on the binomial test (Estim. prob.), and the adjusted p-value corrected for 670 multiple testing using the false discovery rate method (Adj. p-value) is included. For all species, the proportion of juveniles growing associated with other plants is 671 672 significantly greater than expected by chance.

673

Species	Total N.	N. juv.	Estim. Prob.	Adj. p-value
	juveniles	associated		
T. libanitis	203	120	0.591	< 0.001
H. squamatum	539	321	0.596	<0.001
H. syriacim	201	106	0.527	< 0.001
T. moroderi	22	13	0.591	< 0.001
H. stoechas	23	12	0.522	<0.001

674

675

678

679 Figure 1. Experimental design. Fitness components for seedling/juveniles and adults were compared between conspecific plants growing associated with other plants and 680 681 solitary (in brackets). In the field, two different situations were selected, vegetation 682 patches where the focal species was the largest in the patch (i.e. originally acted as the 683 nurse) and where it was an adult but not the largest of the patch (i.e. originally 684 facilitated as a seedling). In the case of seedlings fitness components, the seedling 685 cannot be the largest plant in the patch, and thus this situation is biologically 686 unavailable. We expect that both the originally facilitated and nurse species will 687 benefit (blue) from growing associated compared to solitary, the former at an early 688 developmental stage, increasing its frequency of establishment and growth of 689 juveniles, and the later at a later developmental stage, increasing flower production.

690

Figure 2. Juveniles size when growing in association with other plants or isolated.
Estimated means and standard errors of the height of juveniles (in cm) for plants
growing in vegetation patches and solitary.

694

Figure 3. Benefits of growing in association with other plants in vegetation patches based on the role of the plant (largest plant of the patch assumed to function originally as nurse, and not the largest plant assumed to be originally facilitated as a seedling, see Methods). Estimated means and standard errors of flower production are presented, based on the residuals of the relationship between flower production and height. Benefits were calculated as the difference in flower production between paired individuals growing associated with vegetation patches and solitary.

	Originally nurse	Originally facilitated	
Juveniles: seedling establishment and growth of juveniles	Biologically not posible		
Adults: Reproductive investment	Solitary Solitary	Solitary Growing associated	



Juveniles growth

