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37

38

Abstract

39 Assessing the non-market value of biodiversity conservation is crucial to justify it
40 economically. Using a choice experiment on wetland restoration in Hokkaido, northern
41 Japan, we assessed the willingness of citizens to pay for different ecological statuses of
42 a flagship species (absence, occasional occupancy, permanent occupancy, and breeding)
43 and other principal conservation targets (establishment of a birdwatching station and
44 wetland sizes). The results showed that the fundraising potential of the flagship species
45 surpassed those of other conservation targets, irrespective of its ecological status,
46 highlighting the superior publicity generated by charismatic species. We also showed
47 that upgrading ecological status from occupancy to breeding did not result in additional
48 financial support. Our study emphasizes that, although publicizing ecologically
49 important statuses such as breeding is critical for successful conservation efforts,
50 focusing much effort on flagship species rather than other conservation targets may be
51 important to increase the economic value of conservation practices if such species are
52 available.

53

54 **1. Introduction**

55 Successful biodiversity conservation requires stable and reliable public support because
56 all conservation practices inevitably need persistent budgets (Christie et al. 2006;
57 Addison et al. 2016; Bennett et al. 2016). A promising approach to economically
58 successful conservation is the assessment of non-market values of biodiversity
59 conservation in monetary terms because unravelling these values can help to develop
60 sustainable budget systems (White et al. 2001; Veríssimo et al. 2011; Di Minin et al.
61 2013; Yamaura et al. 2016a). However, most studies still evaluate the effectiveness of
62 conservation practices primarily based on ecological aspects (Iftekhar et al. 2016).

63 Flagship species are species used to raise financial support and public
64 awareness for conservation actions, and it is well appreciated that economic values of
65 flagship species can enhance their conservation as well as broader biodiversity
66 conservation (Caro & O’Doherty 1999; Caro 2010; Veríssimo et al. 2011). Although
67 several studies have shown that the economic values of flagship species vary depending
68 on the population size and public awareness of the species (e.g., Richardson & Loomis
69 2009; Jacobsen et al. 2012; Morse-Jones et al. 2012), the economic value of
70 conservation practices for flagship species with different ecological statuses (e.g.,
71 occupancy or breeding) remains an open question. Unraveling this is critically

72 important to make conservation with flagship species both ecologically and
73 economically sound, because the differences between the ecological statuses can be
74 vital for species' long-term persistence (Donovan & Thompson 2001; Schlaepfer et al.
75 2002; Battin 2004), and loss of species might accelerate the estrangement of people
76 from nature (Miller 2005).

77 Another longstanding issue is the relative value of flagship species compared
78 with other key conservation targets, such as habitat metrics and ecosystem services (but
79 see Kontoleon & Swanson 2003). Metrics such as habitat type and size can be reliable
80 proxies for various components of biological communities, and the importance of
81 ecosystem services has been recognized from both ecological and social aspects
82 (Ferraro & Kiss 2002; Banks-Leite et al. 2011); however, the fundraising potential of
83 these may be relatively poor compared with those of flagship species. This suggests that
84 conservation practices based on these targets may lack sustainable financial support.
85 Therefore, comparisons of monetary values among conservation targets are crucial to
86 conduct economically efficient conservation practices.

87 The present study compared non-market values in monetary terms among
88 conservation practices, focusing on different statuses of flagship species and other
89 principal conservation targets. To achieve this, we conducted a choice experiment (CE)

90 on a wetland restoration project in Hokkaido, northern Japan, and evaluated the
91 willingness to pay (WTP) of citizens for four different statuses (absence, occasional
92 occupancy, permanent occupancy, and breeding) of a flagship species, as well as for
93 restoration with a birdwatching station, varied wetland sizes, and an option to suspend
94 flood-control measures. We used the red-crowned crane *Grus japonensis* as our focal
95 flagship species for two reasons: 1) the name, appearance and habitat of the red-
96 crowned crane are potentially well known by many national citizens, especially those in
97 our study region, because this species is a symbol of good fortune and thus not only a
98 variety of products including Japanese traditional crafts but also several local
99 governments and organizations use it as their symbol (Harris 1994; Wild Bird Society of
100 Japan 2011). This is an important feature for flagship species (Caro & O'Doherty 1999;
101 Caro 2010; Veríssimo et al. 2011). 2) This species has been used to raise funding for
102 conservation actions and environmental awareness for a long time (i.e., the species has
103 been practically used as a flagship species). For example, Wild Bird Society of Japan
104 has established 20 protected areas with the total of 2,516.9 ha in eastern Hokkaido for
105 their habitat protection (Wild Bird Society of Japan 2011), all of which are purchased by
106 donation from members of the society. Based on the results, we discuss how flagship
107 species can be used to raise financial support to conserve ecologically important

108 components and ecosystem services.

109

110 **2. Materials and Methods**

111 **2.1 The flagship species**

112 The red-crowned crane is a large crane species (body length: 140 cm, wingspan: 240
113 cm) distributed across the Korean peninsula, northeastern China, the Russian Far East,
114 and Hokkaido, northern Japan. The population in Hokkaido experienced a drastic
115 decline by the mid-20th century due to widespread replacement of wetlands with
116 agricultural lands. Since then, its population has been increasing gradually due to
117 intensive conservation activities, and its current population size is estimated to be over
118 1,000 individuals. They occur residually in wetlands across Hokkaido, except for in
119 its southern region. The red-crowned crane is considered an endangered species in
120 Japan and is the prefectural bird species of Hokkaido. In addition, our previous study
121 has shown that the species may act as an umbrella species for wetland birds due to its
122 specialized habitat requirements and large home ranges (Higa et al. 2016).

123

124 **2.2 The focal restoration project**

125 To prevent severe damage to agriculture and residents by unpredictable floods, a flood

126 control project was conducted in central Hokkaido, northern Japan (42°55'12"N,
127 141°42'02"E). A total of 200 ha of agricultural land in this area was purchased by a
128 national organization and replaced with a flood-control pond. As a consequence,
129 wetland vegetation such as Manchurian wild rice *Zizania latifolia* has grown, and
130 several wetland animals, such as waterfowl and dragonflies, have started to recolonize
131 the pond. Therefore, the pond is expected to provide suitable habitats for diverse
132 wetland organisms. Currently, no biodiversity maintenance is being conducted in the
133 pond. However, future directions for this project to restore wetlands are under
134 consideration.

135

136 **2.3 The choice experiment to measure willingness to pay**

137 The CE uses a stated preference methodology to estimate values of non-market goods
138 and services (Adamowicz et al. 1998; Louviere et al. 2000). In the CE, we designed a
139 scenario providing a hypothetical flood-control pond with additional features, including
140 non-market values of wetland biodiversity and functions. The CE included the
141 following four attributes (Table 1): 1) the red-crowned crane with four levels
142 (occasional occupancy, permanent occupancy, breeding, and absence). We used these
143 four levels because these statuses may be suitable to represent the habitat quality of a

144 species (Sergio & Newton 2003; Senzaki et al. 2015). 2) A birdwatching station with
145 two levels (establishment and no establishment). The attribute was included to explain
146 the recreational values of the pond with increasing bird abundance or species richness
147 due to restoration plans. This attribute can also separate the value of increased bird
148 abundance due to restoration from the value of the red-crowned crane itself. 3) Wetland
149 sizes with five levels (25, 50, 75, 100, and 125 ha). This attribute was included because
150 habitat area is a key landscape metric for biodiversity conservation, including species
151 richness and abundance (Yamaura et al. 2016b). The levels of this attribute were
152 decided based on the current sizes of natural wetlands around the study area (Senzaki &
153 Yamaura 2016) and the largest wetland size (200 ha) that can realistically be restored in
154 Japan. 4) Annual tax payments needed to introduce preferred plans. For this attribute,
155 we used five levels [100, 500, 1,000, 2,000, or 5,000 JPY (Japanese yen; 100 JPY
156 equals approximately one US dollar)].

157 During times of flooding, the habitat of red-crowned crane and bird-watching
158 station are assumed to be underwater. The option to suspend the flood-control function
159 allows managers to wait to open the floodgates until the last possible moment. This
160 option appeals to beneficiaries of non-market goods and services, but places local
161 residents at a greater risk of flooding. The political conflicts in which this kind of

162 facility becomes embroiled are often related to the fact that the groups who enjoy the
163 benefits are not the same as those who bear the costs (Field 2008). Thus, to measure
164 possibility of conflict between local residents and the general public with respect to
165 tradeoffs between basic flood control services and additional non-market goods and
166 services aforementioned above, the option to suspend the flood-control function with
167 two levels (establishment and no establishment) was also included as an additional
168 attribute.

169 The survey consisted of an introductory text explaining the purpose of the
170 research and the background of the project, the CE scenario, and questions about
171 personal socio-economic characteristics and ecological knowledge (Table 1). We
172 included gender, age and income as the focal personal socio-economic characteristics
173 because these may be related to respondent choices in the CE (Di Minin et al. 2013). In
174 addition, we asked respondents whether they knew the words of “endangered species”
175 and “umbrella species” and have observation experience with the red-crowned crane.
176 These were included because ecological knowledge and natural experience of
177 respondents might affect their willingness to pay (Turpie 2003). In the CE, each
178 respondent evaluated three profiles (alternative management plans) with different levels
179 of the five attributes. We included the status quo profile, which provides no

180 conservation actions (i.e., no crane, no birdwatching station, 25 ha wetlands, no
181 suspension of flood-control function, no annual tax payment), in one of the three
182 profiles. This kind of choice task was repeated seven times for each respondent (i.e.,
183 seven choice sets with different combinations of the levels were used for each
184 respondent). In this study, the profiles were designed using a D-optimal design, which is
185 frequently used in empirical studies (Zwerina et al. 1996). In November 2015, a
186 research company sent invitation emails regarding our Internet questionnaire to its
187 registerees in Hokkaido Prefecture, with taking into account the balance among socio-
188 economic characteristics. Of these registereess, 1,206 (15.7 %) completed the
189 questionnaire. Note that respondents who began the questionnaire but did not complete
190 it were not included in the sample (i.e., there was no incomplete questionnaire in the
191 sample). Thus, all completed questionnaires were used for the analysis. The CE was
192 performed in accordance with relevant guidelines and regulations from the Japanese
193 Ministry of the Environment and were conducted under the current Japanese laws.

194

195 **2.4 Latent class modeling and willingness to pay estimates**

196 Although we used a conditional logit model for the initial analysis (McFadden 1974), its
197 goodness of fit was poor [adjusted McFadden's pseudo $R^2 = 0.031$], suggesting that the

198 preferences of respondents were heterogeneous. Therefore, we used a latent class model
 199 (LCM) for the analysis because it can assess preferences of respondents for attributes in
 200 each of several homogeneous segments. An LCM postulates a random utility framework
 201 consisting of an observable deterministic component and unobservable random
 202 component. Utility for a profile i is described as equation:

203

$$204 \quad U_{ni|k} = \beta'_k X_i + \varepsilon_{ni|k} \quad (1)$$

205

206 where $U_{ni|k}$ is the utility obtained by an individual n in segment k , β'_k is a vector of
 207 parameters of segment k , X_i is a vector of attributes of the i th alternatives, and $\varepsilon_{ni|k}$ is
 208 a vector of the random component assumed to have a type 1 extreme distribution. The
 209 probability that profile i is chosen among a choice set C , is the probability that $U_{ni|k}$ is
 210 larger than any other profile $U_{nj|k}$. The joint probability of a set of seven choices
 211 (i_1, \dots, i_7) from a series of choice sets conditional on belonging to segment k can be
 212 obtained as the following equation:

213

$$214 \quad P_n(i_1, \dots, i_7|k) = \sum_k P_{n|k} \cdot \prod_{l=1}^7 P_{n|k}(i_l|k) =$$

$$215 \quad \sum_k \frac{\exp(\lambda \gamma'_k Z_n)}{\sum_k \exp(\lambda \gamma'_k Z_n)} \cdot \prod_{l=1}^7 \left[\frac{\exp(\mu_k \beta'_k X_{i,l})}{\sum_{j \in C} \exp(\mu_k \beta'_k X_{j,l})} \right] \quad (2)$$

216

217 where μ_k is and a scale parameter for segment k , $\beta'_k X_{j,l}$ is a specific utility for in l th
218 choice set for segment k , and $P_{n|k}$ is a probability that respondent n is classified in
219 segment k . Following Swait (1994) and Boxall and Adamowicz (2002), we assumed a
220 latent membership likelihood function represented by psychometric or socioeconomic
221 characteristics of respondent n . The probability is described by a scale parameter (λ),
222 psychometric or socio-economic characteristics of a respondent (Z_n) and their
223 parameters (γ'_k), and a vector of the random component. In contrast to the conditional
224 logit model, LCM assumes that an individual n belongs to a latent class k that is
225 unobservable a priori. The model is estimated based on maximum likelihood estimation.
226 We also estimated attribute-specific WTP using the parameters obtained by the LCM
227 based on the following equation:

228

$$229 \quad \text{WTP}_a = -\frac{\beta_{ak}}{\beta_{taxk}} \quad (3)$$

230

231 where β_{ak} is a segment-specific parameter (non-monetary coefficient) of each attribute
232 and β_{taxk} is the segment-specific parameter of additional tax payment.

233

234 **3. Results**

235 The LCM analysis required determining the number of segments. We initially sought
236 the number of segments based on two major statistical criteria (AIC, BIC). However,
237 these procedures required too many number of segments to interpret (i.e., more than
238 four segments) and hence made a characteristic of each segment obscure. In such cases,
239 a model selection can be conducted based on the use of theory and common sense
240 (Agresti 2002). Thus, we have decided to use two segments because the two-segment
241 LCM enabled us to discuss which socio-economic variables and ecological knowledge
242 were related to WTP (see the following results and Table 2).

243 The LCM with two segments greatly improved model fit (adjusted McFadden's
244 pseudo $R^2 = 0.302$) over the conditional logit model (Table 2). The first segment
245 constituted 41.9% of the sample and the socio-economic characteristics of this segment
246 (i.e., age, gender and income) did not differ from those of the second segment (Table 2).
247 Respondents in this segment did not know the words “endangered species” and
248 “umbrella species”, had no observation experience with the red-crowned crane, and had
249 no significant interest in the status of the red-crowned crane (Table 2). In addition, the
250 alternative-specific constant for the status quo profile was positive and its absolute
251 value was quite large compared with other parameters, indicating that they preferred the

252 current situation (Table 2). The WTP for the birdwatching station was estimated to be
253 76.12 JPY (−76.12 JPY is the baseline level that provides no birdwatching station),.

254 The second segment constituted 58.1% of the sample (Table 2). Compared to
255 the first segment, respondents in this segment significantly knew the words “endangered
256 species” and “umbrella species”, and had observation experience with the red-crowned
257 crane (Table 2). In addition, all parameters except for the wetland size attribute were
258 significantly different from zero at the 5% level (Table 2). The parameter for the status
259 quo profile was negative, indicating that respondents expected a conservation plan. The
260 crane attribute had its greatest increase in WTP from absence (as a baseline level of
261 effects-coded dummy variables; −1,756.45 JPY) to occasional occupancy (381.73 JPY,
262 a difference of 2,138.18 JPY; Fig 1). The differences in WTP among ecological statuses
263 other than absence were relatively small (occasional to permanent occupancy: 251.80
264 JPY, permanent occupancy to breeding: 107.65 JPY; Fig. 1). WTP of other attributes
265 were 390.11 JPY for a birdwatching station (−390.11 JPY is the baseline level that
266 provides no birdwatching station), and 205.38 JPY for the option to suspend flood-
267 control functions (−205.38 JPY is the baseline level that provides no such option).
268 These results suggest that the economic values of these attributes would be considerably
269 lower than those of the crane (Fig. 1).

270

271 **4. Discussion**

272 Currently, neither the fundraising potential of different ecological statuses (e.g.,
273 occupancy or breeding) of flagship species nor their economic values relative to other
274 principal conservation targets have been quantified, despite their clear importance to
275 conservation practices. In this study, 41.9% of respondents who had no ecological
276 knowledge and observation experience relevant to a flagship species (the red-crowned
277 crane) were less interested in implementation of any conservation practices. In contrast,
278 the other 59.1 % respondents who had excellent ecological knowledge and observation
279 experience relevant to the flagship species valued conservation practices with the
280 flagship species more than those without such species (Fig. 1). These results suggest
281 that facilitating knowledge and experience with flagship species, for people categorized
282 to the former segment in particular, may be important for increasing financial support,
283 although we cannot exclude the possibility that there are other superior attributes such
284 as other species. Additionally, we showed that consideration for ecologically important
285 statuses of the flagship species did not always improve the economic value of
286 conservation practices.

287 The higher WTP for the red-crowned crane compared with other attributes

288 could be explained by the red-crowned crane's high-profile status and publicity, since it
289 is considered an umbrella, symbolic and endangered bird species in Hokkaido (Higa et
290 al. 2016). This explanation was also supported by lower WTP for a birdwatching
291 station, as this suggests that people were less interested in other bird species in the
292 wetland. Therefore, our results emphasize the strong potential of "charismatic species,"
293 regardless of ecological status, as financial drivers of conservation practices in at least
294 our study region, which was consistent with previous studies (e.g., White et al. 1997;
295 White et al. 2001; Di Minin et al. 2013).

296 The small differences in WTP among the ecological statuses other than absence
297 of the flagship species suggest that ecological status does not foster additional financial
298 support. This is unfortunate for two reasons. First, generally speaking, a status of
299 permanent occupancy or breeding can be an excellent indicator of population
300 persistence (Sergio & Newton 2003). Second, conservation of a single species based on
301 their breeding status could be a shortcut to identify productive areas of many co-
302 occurring species (Senzaki et al. 2015). Although future research should examine the
303 reasons why people did not value ecologically important status, a potential reason for
304 this result is that respondents might not recognize the differences among the ecological
305 statuses. Therefore, our study indicates that marketing for raising the publicity of

306 ecologically important statuses may be important. To do so, enhancing interactions
307 between ecologists and citizens through blogs, newsletters, or workshops may be
308 promising solutions (Dicks et al. 2014; Matzek et al. 2014). It may also be critical to
309 explore when, where, and how ecologists can better educate the public on the
310 importance of ecological status, highlighting an urgent need for better collaboration
311 between conservation and social sciences (Martín-López et al. 2008; Soga et al. 2016).

312 The birdwatching station had lower WTP and wetland sizes had no WTP,
313 indicating that conservation practices using landscape metrics and ecosystem services as
314 targets may be prone to lacking sustainable financial support. The absence of WTP for
315 wetland size is important because habitat area is considered a key metric in many
316 conservation practices, including restoration and setting of priority areas (e.g., Faith
317 2003). One potential solution to this is a hybrid approach that uses several non-mutually
318 exclusive targets simultaneously. For example, even if management is focused primarily
319 on recovering natural habitats or ecosystem functions, using a flagship species may
320 raise its monetary value. This hybrid approach has great potential because flagship
321 species with high fundraising potential can exist among even common species (Smith et
322 al. 2012). In addition, specific attributes of flagship species related to individual
323 preferences and financial support are also revealed (Martín-López et al. 2008; Veríssimo

324 et al. 2014a, b).

325 Finally, it must not be forgotten that the WTP for the option to suspend flood-
326 control functions in the larger group was positive and significant. The results show that
327 this wetland restoration project is generally favorable to the public. However, the
328 operation of the flood-control pond may stoke conflicts between the general public and
329 local residents. Both the amount of financial support and its distribution (e.g.,
330 compensation for local residents who bear a burden caused by additional services)
331 require further discussion. In addition, although our results were based on respondents
332 with diverse socio-economic characteristics throughout our study area and thus might be
333 considered generalizable, the response rate (15.7 %) was not high. Thus, future studies
334 should follow up the generality of our findings.

335 In conclusion, our study showed that the flagship species in this study,
336 irrespective of their ecological status, was more valued than other important
337 conservation concepts such as ecosystem services and habitat metrics. This suggests that
338 a single flagship species can play an important role in raising the economic value of
339 conservation practices. Although raising the public awareness of ecologically important
340 statuses may be crucial to justify conservation practices both ecologically and
341 economically, our results suggest that focusing much effort on flagship species rather

342 than other conservation targets may be important to increase the economic value of
343 conservation practices if such species are available.

344

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350

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469

470 **Table 1.** Attributes and corresponding levels used in the choice experiment and socio-economic and ecological knowledge questions. In
 471 the CE, we presented three profiles with different levels of the five attributes to each respondent and repeated this kind of choice task for
 472 seven times.

Attributes	Levels
Status of the red-crowned crane	Occasional occupancy/Permanent occupancy/Breeding/Absence
Wetland area	25ha/50ha/75ha/100ha/125ha
Bird-watching station	Yes/No
Option to enable suspension of flood-control function	Yes/No
Annual amount of tax payment for a restoration plan	100JPY/500JPY/1,000JPY/2,000JPY/5,000JPY
Observation experience of the red-crowned crane	Yes/No
Knowledge of endangered species	Yes/No
Knowledge of umbrella species	Yes/No
Gender	Male/Female
Age	10~19, 20~29, 30~39, 40~49, 50~59, 60~69, 70~
Income	~2,000,000JPY/~4,000,000JPY/~6,000,000JPY/~8,000,000JPY/~10,000,000JPY/~12,000,000JPY/~14,000,000JPY/~16,000,000JPY/~18,000,000JPY/~20,000,000JPY/~22,000,000JPY/22,000,000JPY~

473

474 **Table 2.** Results of latent class model parameters for each recognized segment.

Attributes	Levels	Segment_1			Segment_2		
		Coef	SE	<i>p</i> -value	Coef	SE	<i>p</i> -value
Status of the red-crowned crane [†]	Occasional occupancy	0.20	0.12	0.09	0.18	0.04	<0.001
	Permanent occupancy	0.02	0.12	0.87	0.30	0.04	<0.001
	Breeding	0.03	0.14	0.83	0.35	0.04	<0.001
	Absent	-0.25			-0.83		
Wetland area (*10 ⁻²)		0.37	0.19	0.053	0.09	0.05	0.067
Bird-watching station [†]	Yes	0.19	0.09	0.038	0.19	0.02	<0.001
	No	-0.19			-0.19		
Option to enable suspension of flood-control function [†]	Yes	-0.10	0.08	0.22	0.10	0.02	<0.001
	No	-0.10			-0.10		
Annual amount of tax payment for a restoration plan (*10 ⁻³)		-2.45	0.24	<0.001	-0.47	0.02	<0.001
Alternative-specific constant for status quo profile		1.91	0.30	<0.001	-0.76	0.09	<0.001
Constant term of membership function		0.18	0.15	0.21	0.00 ^{††}	-	-
Crane experience		-0.36	0.13	0.006	0.00 ^{††}	-	-
Knowledge of endangered species		-0.45	0.13	<0.001	0.00 ^{††}	-	-
Knowledge of umbrella species		-1.22	0.40	0.002	0.00 ^{††}	-	-

Sex	0.13	0.12	0.31	0.00 ^{††}	-	-
Age	-0.07	0.12	0.60	0.00 ^{††}	-	-
Income	-0.002	0.001	0.24	0.00 ^{††}	-	-
Class probability		0.42				0.58
Log likelihood						-6460.37
Adjusted McFadden's Pseudo R^2						0.302
No. of observations						8442

[†] These attributes are included in the model as dummy variables with effects coding (Louviere et al., 2000).

^{††} These are fixed parameters.

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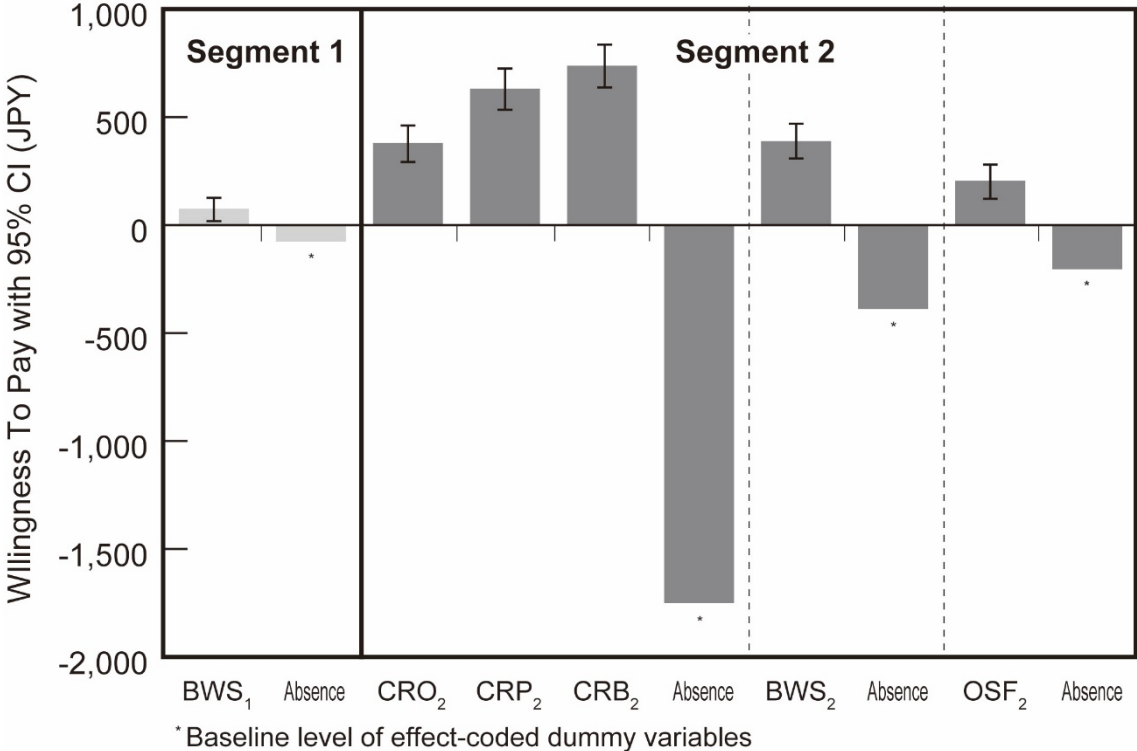
477 **Figure legends:**

478 **Figure 1** Willingness to pay (WTP) for choice experiment attributes derived from the
479 latent class model. Error bars in the upper panel are 95% confidence intervals calculated
480 from the coefficients and variance terms of the latent class model. “CRO,” “CRP,”
481 “CRB,” “BWS,” and “OSF” indicate “occasional occupancy of the crane”, “persistent
482 occupancy of the crane”, “breeding of the crane”, “bird-watching station”, and “option
483 to suspend the flood control function,” respectively.

484

485

486 **Figure 1**



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