1 High nitrogen rates do not increase canola yield and may affect soil bacterial functioning F Becker¹, C MacLaren^{1, 2}, CJ Brink³, K Jacobs³, MR le Roux¹, PA Swanepoel^{1*} 2 ¹Department of Agronomy; Stellenbosch University 3 4 ²Sustainable Agricultural Sciences, Rothamsted Research ³Department of Microbiology, Stellenbosch University 5 6 *Corresponding Author: pieterswanepoel@sun.ac.za; +27218084668 **ORCID IDs** 7 CM: 0000-0002-6700-3754; KJ: 0000-0003-3972-5343; CB: 0000-0003-2607-6831; MR: 0000-8 9 0002-6473-237X PS: 0000-0002-6481-0673 10 Author contributions: PS and MR designed the study. FB undertook fieldwork, and FB and CB 11 12 completed the laboratory tests. CM analysed the data, and FB, CM and PS wrote the article with contributions from CB and KJ. All authors provided critical revisions to the final version. 13 14 **Core ideas** 15 • Over-fertilization of N is common in canola in South Africa 16 • We evaluated N fertilization rates and application timings for canola 17 • Canola growth and yield were generally not affected by N rates or timing 18 • A low N rate favoured soil bacterial communities more likely to mineralize N 19 20 • A low N rate (60 to 90 kg ha⁻¹) split into two or three applications can sustain yield and have minimal effects on soil bacterial functioning 21 22 **Abbreviations**: ARISA = Automated Ribosomal Intergenic Spacer Analysis; CLPP = Community 23 level physiological profiling: carbon source utilization; DAE = days after emergence; LAI = leaf 24 area index; NMS = non-metric multidimensional scaling; OTU = operational taxonomic unit 25

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26 Abstract

Nitrogen fertilization has a fundamental role in agricultural productivity. However, injudicious N 27 applications to crops are common. It is important to ensure the minimum N required for 28 satisfactory crop growth is applied but that excess amounts are avoided due to potential impacts 29 on agroecosystem functioning. Nitrogen at 0, 60, and 150 kg ha⁻¹ was applied as limestone 30 ammonium nitrate to plots arranged in a randomized complete block design, on three farms to 31 32 determine the impact of rate and temporal distribution of fertilizer on canola (*Brassica napus* L.) production in South Africa, and the effect of N fertilizer application on the composition and 33 34 diversity of soil bacterial communities. The amount and distribution of N had only minor effects on canola growth (P < 0.05) and no effects on yield or harvest index. Splitting fertilizer into two 35 or three applications throughout the season resulted in more mineral N available in the soil later 36 37 in the season. Increasing the N rate from 60 to 150 kg ha⁻¹ had a significant impact on bacterial community composition. The lower rate favored bacteria that are more able to break down N-38 containing carbon sources. No effects of fertilizer amount or distribution were observed on either 39 40 N fixation potential (number of nifH gene copies) or bacterial community diversity. Overall, a low rate of N fertilizer split into multiple applications is recommended for canola production, as 41 higher rates do not increase yield and may have a detrimental impact on soil carbon and nitrogen 42 cycling. 43

44 Introduction

The need for sustainable agricultural production systems is well recognized, with substantial 45 increases in the human population posing huge challenges for future agricultural production within 46 47 Earth's environmental limits (Tilman et al., 2002; Rockström et al., 2017). Environmental degradation associated with agriculture is exacerbated by external inputs, in particular with 48 injudicious inorganic fertilizer applications. Inorganic N fertilizer application is a primary 49 50 approach to intensify crop production and ensure food security (Liu et al., 2011), but it also has a large C footprint and causes eutrophication and algal blooms when it leaches from agricultural 51 52 land to waterways (Seitzinger and Phillips, 2017). In addition, fuel and energy use during the manufacturing and transport processes of fertilizers may also contribute substantially to 53 54 environmental erosion (Shibata et al., 2017). Although N fertilization has a fundamental role in 55 agricultural productivity, it is important to ensure the minimum N required for satisfactory crop 56 growth is applied.

57 Canola (Brassica napus) was introduced into crop rotation systems of South Africa in 1994 58 to increase crop diversity (BFAP 2018). Canola production in South Africa is growing, and according to predictions made by the Bureau for Food and Agricultural Policy (BFAP 2018), will 59 increase to ca. 200,000 tons by 2027. Canola has a higher N demand than most other cash crops 60 (Ma and Herath, 2015), but dependence on inorganic N fertilizers to increase canola production is 61 62 not sustainable and, therefore, attention should be shifted to retaining and fixing N on-farm through 63 optimising soil biological activity. Moreover, due to the lack of robust guidelines tailored for the South African climate and soil conditions, N fertilization is often applied injudiciously, which may 64 have negative impacts on the finely balanced interactions in the soil environment. There is a need 65 66 to understand how much inorganic N is necessary to produce satisfactory canola yields in the region, and how different fertilization strategies affect soil biological function. 67

68 The relationship between soil microbial diversity, ecosystem functioning, associated services, and management practices (e.g. N fertilization) is under increasing scrutiny to elucidate 69 the complexities that underpin the productivity of agroecosystems (Brussaard et al., 2007; 70 71 Hartmann et al., 2015; Hartman et al., 2018). Increased biodiversity in the microbial community may enhance the functional capacity of the soil ecosystem (Bender et al., 2016). Thus, the use of 72 agricultural management practices that can maintain soil functional diversity is advocated in order 73 74 to build inherent resilience to environmental shocks. The same motive has driven the 75 implementation of crop rotation systems in the Western Cape (Venter et al., 2017).

76 It is widely acknowledged that N fertilizer may affect soil biodiversity and on-farm ecological functioning, potentially decimating the ecosystem services provided by the soil 77 microbial community (Bisset et al. 2011; Gordon et al., 2016; Hartmann et al., 2015; Jackson et 78 79 al., 2017). For example, the soil N and C cycles are mediated by soil bacteria, which are involved 80 in the build-up and decomposition of soil organic matter (Jackson et al., 2017), and in the conversion of N between its organic and inorganic forms (Kuypers et al., 2018). It has been shown 81 82 that microbial communities can be sensitive to fertilizers, particularly at the rates applied in current agroecosystems (Gordon et al., 2016; Hartmann et al., 2015). This suggests that N fertilizer 83 applications may interfere with the capacity of the microbial community (including beneficial 84 bacteria) to cycle N and C in ways that are beneficial to both crop growth and C sequestration. 85 86 Thus, optimising N fertilizer applications should have benefits for both protecting the off-farm 87 environment and sustaining the capacity of farm soils to produce crops.

Our understanding of the effect of N fertilization of canola on soil bacterial communities is currently limited. The overall aim of this study was to evaluate different N fertilization rates and N distribution for canola production in South Africa, but it specifically seeks to determine the effect of fertilizer N application on the composition and diversity of soil bacterial communities. The relationships between different N fertilizer application strategies, the soil bacterial Page 5 of 36

community, and canola growth and yield were assessed. The effects of both the quantity of N
fertilizer and whether it was all applied at once or distributed throughout the season (at planting
and as a top-dressing) were investigated.

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97 Materials and methods

98 Trial location

99 Trials were conducted in 2016 at three farms in the winter rainfall area within South Africa's Western Cape province: Langgewens Research Farm (33°16'36.6"S, 18°42'11.4"E), Roodebloem 100 Experimental Farm (34°13'29.5"S 19°31'47.3"E) and Altona, a commercial farm (33°42'15.6"S, 101 18°38'12.3"E). Langgewens and Altona are located in the Swartland region, and typically receive 102 440 mm and 690 mm of rain per year with 85-90 % of rainfall falling in the colder winter months 103 104 (April-September). Roodebloem is located in the southern Cape region and receives 585 mm of 105 rain per year, of which 80 % falls in the winter months. In 2016, annual rainfall and temperature 106 patterns were similar to long-term averages, with the exception that May was unusually dry 107 (records were obtained from weather stations either on or nearby each farm). Soils on Roodebloem Experimental Farm are generally shallow (<400 mm deep), shale-derived soils of a sandy loam 108 texture. The parent material of soils in the Altona and Langgewens districts are mainly derived 109 from greywacke and phyllite with limited pedological development, therefore shallow (<400 mm 110 deep). The soil chemical and physical characteristics of each site is presented in Table 1. 111

112

113 Experimental design

The trials followed a crossed full factorial design, with treatments receiving either 60 kg ha⁻¹ or 150 kg ha⁻¹ of N, of which 20 kg ha⁻¹ was applied at planting and the remainder distributed in either one, two or three applications later in the season (at 30, 60 and 90 days after emergence; DAE). A null control was also included, which received no N at any point in the season, so the trial included seven treatments in total (Table 2). These were laid out in a randomized complete block design, with four replicates at each of the three farms. Plots were 2.75 x 5 m. Half of each plot was intended for destructive measurements (sampling of plants), while the other half was used for yield determination. Measures of canola plant production and soil bacterial community composition and function were taken in each plot at various time points throughout the season (Table 3). The methods for each of these are detailed in the following sections.

124

125 Trial management

126 Weeds were eradicated prior to planting with paraquat. A fine seedbed was created using a 21 tine vibro flex to a depth of 150 mm, during which the pre-emergence herbicide Trifluralin [2,6-127 Dinitro-N,N-dipropyl-4-(trifluoromethyl)aniline] was applied so that it could be incorporated into 128 129 the soil. The insecticide chlorpyrifos (O, O-diethyl O-3, 5, 6-trichloropyridin-2-yl 130 phosphorothioate) was applied just before planting. Phosphorous, in the form of double superphosphate, was applied at a rate of 20 kg ha⁻¹ on the day of planting, according to 131 132 recommendations from soil tests done prior to establishment of the trial. No potassium was required according to soil test results (Table 1). Nitrogen was applied in the form of limestone 133 ammonium nitrate (LAN) as specified in Table 3. 134

Canola was sown using a Wintersteiger disc plot planter with 170 mm row spacing. The triazine tolerant canola cultivar, Hyola 555 TT was planted at 4 kg ha⁻¹ on 9 May 2016 at Langgewens, 5 May 2016 at Altona, and 4 May 2016 at Roodebloem. The preceding crops on Langgewens, Altona and Roodebloem were wheat (*Triticum aestivum*), annual medics (*Medicago* spp.), and oats (*Avena sativa*), respectively. Methiocarb [(3,5-dimethyl-4-methylsulfanylphenyl) N-methylcarbamate] was applied just after planting to control snails, slugs and millipedes. At 30 DAE, atrazine (6-chloro-4-N-ethyl-2-N-propan-2-yl-1,3,5-triazine-2,4-diamine) and chlorpyrifos

- were applied to control grass weeds and insects, respectively. Methiocarb was also applied at 30DAE.
- 144

145 Soil nitrogen content analysis

Soil cores (ø 45 mm) were taken to a depth of 150 mm. Three sub-samples were taken from each 146 plot and combined to form a single composite sample per plot, then air-dried at room temperature, 147 and sieved with a 1 mm sieve. Samples were taken 30, 60, 90 DAE and at when the canola reached 148 physiological maturity (approx. 150 DAE). Soil samples were analysed for ammonium and nitrate 149 150 content using the indophenol-blue (Keeney et al., 1982) and salicylic acid methods (Cataldo et al., 1975) respectively. Total soil mineral N (kg ha⁻¹) was calculated as ammonium plus nitrate, which 151 is the N readily available to plants. To convert the total mineral N concentration (mg kg⁻¹) to stock 152 (kg ha⁻¹) a bulk density of 1400 kg m⁻³ was used, which is the average bulk density for the region 153 reported by (de Clercq et al., 2013). 154

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156 Soil bacterial community analysis

Soil cores (ø 45 mm) were taken to a depth of 150 mm using a stainless steel pipe and a hammer.
The pipe was washed and sterlized with 70 % ethanol between sampling of different plots. Samples
were collected at canola physiological maturity only.

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161 Automated Ribosomal Intergenic Spacer Analysis (ARISA)

The bacterial community composition within each sample was determined with Automated
Ribosomal Intergenic Spacer Analysis (ARISA) (Ranjard et al., 2001). DNA was extracted from
0.25 g of soil using the Zymo research soil microbe DNA MicroPrep[™] kit (Zymo research USA).
Extracted and purified DNA was separated on a 1 % agarose gel stained with ethidium bromide to
visualize under ultraviolet light.

The polymerase chain reaction (PCR) reactions were performed on the purified DNA using 167 ITSReub (5'-GTCGTAACAAGGTAGCCGTA-3') and FAM (carboxy-fluorescein) labelled 168 ITSF (5'-GCCAAGGCATCCACC-3') primer set for the 16S rRNA intergenic spacer region to 169 170 determine bacterial diversity using ARISA (Cardinale et al. 2004, Slabbert et al., 2010b). PCR reactions were done using a 2720 Thermal Cycler (Applied Biosystems, USA). The reaction 171 mixture contained 0.5 µl purified genomic DNA, 500 nM of each primer, 4.1 µl PCR grade water 172 (nuclease free) and 5 µl KapaTaq readymix (Kapa Biosystems, South Africa) for a total volume 173 174 of 10 µl. The PCR consisted of an initial denaturing step of 5 minutes at 94°C, followed by 40 175 cycles at 94°C for 45 s, 56°C for 50 s and 72°C for 70 s. The reaction was completed with a final extension at 72°C for 7 minutes and then cooled and held at 4°C. All the samples were done in 176 triplicate and pooled to compensate for PCR bias. 177

178 The PCR products of the pooled samples were run on an ABI 3010xl Genetic Analyser to obtain an electropherogram of different fragment lengths and fluorescent intensities. Bacterial 179 180 ARISA samples were run along the LIZ 1200 size standard which is designed for sizing DNA 181 fragments in the 100 – 1200 base pair (bp) range. Fluorescence intensities were converted to electropherograms using the Genemapper 5 software. The peaks on the electropherogram represent 182 different fragments of different sizes, termed operational taxonomic units (OTUs), and the heights 183 of the peaks indicate relative abundance of the fragments. The lengths were calculated by plotting 184 185 a best fit curve using the size standard and extrapolating the fragment size from the sample. Only 186 fragment sizes between 100 and 1000 base pairs and peak heights above 150 fluorescent units were used for analysis as OTU's. A bin size of 3 bp was used to minimize inaccuracies of the ARISA 187 profile (Brown et al., 2005; Slabbert et al., 2010b). 188

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190 *Community level physiological profiling (CLPP): carbon source utilization*

The CLPP was done by determining the carbon source utilization of the soil bacterial community. Soil samples were diluted in distilled nuclease free water and inoculated, in triplicate, into Biolog EcoPlatesTM (Biolog Inc., USA). The plates contain 31 different C sources in different wells and a control well containing no C source. Plates were incubated at 28°C. Utilization of the C sources by microbial populations reduce the tetrazolium dye inside the plate wells that cause a colour change. This colour change was measured twice daily over a period of 5 to 10 days with a spectrophotometer at 590 nm to determine the average well colour development (AWCD).

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199 Nitrogen fixation capacity: number of nifH gene copies

The *nifH* PCR product from a *Burkholderia* strain was used for preparation of the standards as 200 201 well as a positive control. The PCR product was purified using GeneJET PCR Purification Kit 202 (Thermo Scientific). The PCR product size of 380 bp was verified by electrophoresis on a 1 % agarose gel. The purified PCR products were quantified using a µLITE (Biodrop, Cambridge, UK) 203 204 and the *nifH* gene copy number was determined using the fragment length, molecular weight and 205 Avogadro's number. The known concentration of the PCR product was used to prepare a standard curve, in triplicate, to measure *nifH* gene copy numbers. The *nifH* gene copy numbers of the soil 206 207 samples were quantified by using quantitative PCR (qPCR) using the nifH F1 and nifH 438r primers (Boulygina et al., 2002; De Meyer et al., 2011). The qPCR assays were performed using 208 209 the LightCycler 96 (Roche) with a SYBR Green 1 fluoroprobe as the protocol suggested (Brink et 210 al., 2019).

A standard curve was generated for every qPCR run ranging from $1x10^{10}$ to $1x10^{0}$ gene copies μ L⁻¹. Each run also included a positive control as well as a negative control. The same soil DNA samples used for ARISA was used for qPCR and was run in duplicate. The reaction volume contained 3 μ L nuclease free water, 2 μ L nifH F1 (100 nM) and nifH 438r (100 nM) primers, 10 μ L SYBR Green I Master Mix (2x) and 5 μ L sample DNA as described by manufacturers specifications. The thermal cycle used for qPCR consisted of 95°C for 5 minutes followed by 40 cycles of 94°C for 60 s, 60°C for 60 s and 72°C for 30 s. A melt curve analysis was done after the 40 cycles to verify specificity of amplicons. This analysis identified the number of nifH gene copies per 5 μ L of DNA, which was converted to the copy number per gram soil based on 100 μ L DNA per 0.25 g soil.

221

222 Canola growth and yield determination

223 Plant population was determined by counting seedlings within the border of a half square meter 224 quadrat at 30 DAE and converted to plants m⁻². Biomass was determined by cutting 10 plants per plot at ground level at 30, 60, 90 DAE and at physiological maturity. The plants were dried in an 225 226 oven at 70°C for 48 hours and weighed. Biomass per plant was converted to biomass m⁻² by using 227 the particular plot's plant population. An additional ten plants per plot were sampled to determine 228 leaf area index (LAI). Leaf area index was measured at 60 and 90 DAE using a LI-COR 3100 leaf 229 area meter. The LAI describes the potential surface area of leaves available for capturing light and 230 thus photosynthetic capacity, and so a higher LAI gives a plant the capacity for higher biomass 231 accumulation and yield potential (Viña et al., 2011).

Canola seed was harvested at physiological maturity on 7 November 2016 at Langgewens,
 9 November 2016 at Altona and 4 November 2016 at Roodebloem with a Hege plot harvester. The
 harvested seed were cleaned by using sieves and weighed to determine the yield per plot (ton ha⁻¹). The harvest index, or proportion of aboveground biomass, was also calculated:

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Ten plants per plot were dried, ground and passed through a 1 mm sieve, then analysed for % Ncontent using the Kjeldahl method (AOAC, 2000).

241

242 Data analysis

Prior to analyses, OTU heights were normalized to the lowest height total before analysis, by
dividing each value within the sample by the total height representing relative abundance of the
DNA fragments, and multiplying each value with the lowest height total (so that the number of
OTUs is underestimated rather than overestimated) (Slabbert et al 2010b) . The OTU Shannon
diversity was calculated according to the following formula:

$$-\sum_{i} p_{i} ln p_{i}$$

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where p is the proportion of biomass in species *i*, and *ln* is the natural logarithm. The Shannon index is an acceptable diversity measure for OTU data (Hill et al 2003).

252 Linear mixed regression models were used to assess differences in soil mineral N, canola 253 biomass, leaf area index, yield, harvest index, bacterial community diversity, and bacterial N fixation capacity. The amount and distribution of N and their interaction were set as fixed effects, 254 while replicate nested in farm was set as a random effect. Farm was included as a random effect 255 256 to explore whether there was any independent effect of the N fertilizer treatments when site was accounted for. Site itself was not a factor of interest, as sites were selected for logistical reasons 257 and not due to any particular characteristics nor prior knowledge of typical canola yields or 258 259 microbial communities. The negative control was included by structuring the model to test for the 260 difference between the control and all treated plots, and to test for differences between treatments 261 nested within all treated plots. A log transformation was used for the response variables of soil mineral N, canola biomass, and LAI so that the data fit the assumptions of linear regression of 262 263 normality and homoscedastic variance in the model residuals.

To assess bacterial community composition, NMDS ordination based on the Bray-Curtis dissimilarity measure was used for both OTUs and carbon sources utilized. PERMANOVA was used to test for significant differences in composition between different amounts and distributions of N with farm as a grouping variable, to detect effects of these variables within potentially different bacterial communities on different farms. It is currently not possible with available software to nest grouping variables in PERMANOVA, otherwise replicate could have been used as a grouping variable too. Data analysis was undertaken in R, version 3.4.3 (R Core Team, 2017).

271

272 **Results**

273 Soil mineral nitrogen and canola production

As the season progressed, soil mineral N tended to become higher in plots that received greater amounts of N applied in a more even distribution (Table 4, Figure 1). Applying 60 kg ha⁻¹ of N across three applications resulted in more available N later in the season than applying 150 kg ha⁻¹ ¹ at 30 DAE in a single dose (Figure 1).

Canola biomass and LAI responded slightly to the amount of N applied and its distribution 278 279 (Table 4), but were generally not different between treatments (Figure 1). When N fertilizer was 280 distributed in three applications compared to one or two applications, biomass and LAI tended to be lower at 60 and 90 DAE for 60 kg ha⁻¹ treatments, but higher in the 150 kg ha⁻¹ treatments at 281 60 and 90 DAE. It is possible that treatment 60(3) did not receive sufficient N for full growth prior 282 to 90 DAE, whereas a more even distribution of fertilizer was advantageous where more N was 283 284 applied in treatment 150(3). However, these small differences in plant growth did not result in any 285 significant differences in either canola harvest index or plant tissue N content at physiological maturity between the treatments, and only the control differed (P < 0.05) from the other treatments 286 (Table 5, Figure 2). In terms of random effects, variability between replicates tended to be very 287 288 low, while variability within plots of the same treatment at the same site (the residual error term) was often similar or larger than the variability between farms (Tables 3 and 4), indicating relatively 289 290 high within-site and within-replicate variability that could not be explained by the treatments.

The lack of substantial differences in canola growth and yield indicates that uptake of N by canola does not explain the reduced soil mineral N observed later in the season in treatments where all N fertilizer was applied early (Figure 1). It can therefore be assumed that the N is lost from the system (perhaps to leaching, bacterial immobilisation or weeds) and not incorporated into crop plant matter.

296

297 Bacterial community composition, diversity and function

298 No significant differences were observed in OTU richness, Shannon diversity or the number of 299 *nifH* gene copies between treatments, and again variability within treatments and replicates (the residual) was higher than variability between replicates and farms (Table 6). However, the non-300 301 metric multidimensional scaling (NMS) ordination (Figure 3) and PERMANOVA of the OTUs 302 found in each plot indicate that N amount does have a significant impact on bacterial community composition (Table 7). In the ordination, plots that received 150 kg ha⁻¹ N are shifted higher along 303 the first axis of the NMS compared to control plots or plots receiving 60 kg ha⁻¹ N at the same 304 farm (Figure 4). This suggests that increasing N levels favours a different group of bacteria to 305 306 those found under low N levels.

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The NMS ordination and PERMANOVA for carbon source utilization suggest that the amount of 308 309 N fertilizer applied may alter microbial function (Table 7, Figure 4). Plots that received more N 310 tend to be shifted higher along axes 1 and 2 of the NMS, which is associated with higher use of 311 carbohydrates, phosphorylated compounds and carboxylic acids and lower use of amino acids, amines, esters and polymers. Amines and amino acids always contain N, while some esters and 312 313 polymers do, so it is possible that lower mineral N availability favours bacteria that can break down N-containing carbon sources and thus extract N as well as C from those sources, to meet 314 315 their N needs. In particular, Figure 4(a) shows a greater association of plots receiving 150 kg N

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- ha⁻¹ with non-N-containing carbon sources, and Figure 4(b) suggests a higher affinity for polymers
 of control plots and plots receiving 60 kg N ha⁻¹.
- 318
- 319 Discussion

320 *Optimising fertilization strategy for canola growth*

321 This study provided no evidence that increasing N fertilizer from 60 kg ha⁻¹ to 150 kg ha⁻¹ has any 322 benefits for canola production in the winter rainfall region of South Africa's Western Cape. There were no detectable differences in yield or harvest index, and only minor differences in plant 323 324 biomass and LAI (Figures 1 and 2). This finding concurs with other recent and ongoing trials in 325 the region, and may be a result of a regional switch to conservation agriculture practices over the last twenty years. Conservation agriculture practices including reduced tillage, maintenance of 326 327 crop residues, and crop rotation can increase soil organic matter and thus increase rates of N mineralization (Plaza et al., 2013; McDaniel et al., 2014), so it is possible these practices may have 328 329 reduced canola N requirements over time. Current fertilizer guidelines for canola in the Western Cape were based on conventional tillage systems. These were determined the from target yield 330 331 potential of canola for specific regions (a result of in-season rainfall), taking soil texture and crop 332 rotation into consideration (Fertasa, 2016). However, preliminary work to re-evaluate fertilizer 333 guidelines for conservation agriculture systems indicates that total seasonal N rates above 50 to 75 kg ha⁻¹ would not be recommended due to the risk of leaching and low N use efficiencies of canola 334 335 (du Toit, 2018).

336 Splitting the fertilizer into two or three applications throughout the season was 337 advantageous compared to applying all fertilizer by 30 DAE, in terms of reducing N loss to the 338 environment. It did not result in yield differences in our study, but other studies have observed a 339 higher canola yield quality (increased oil percentage) when N availability is higher later in the 340 season (du Toit, 2018; Swanepoel et al., 2019). The reduced levels of N observed later in the 341 season under a single application was not compensated for by an increase in crop growth in these 342 treatments, indicating that the N has moved elsewhere. Some N may have been immobilized by microbial activity, but leaching is also likely as N in the form of nitrate is readily soluble in water 343 and thus carried away through rainwater infiltration or surface water run-off. This can have 344 345 detrimental consequences in particular for surrounding aquatic ecosystems, with fertilizer run-off 346 a major contributor to eutrophication and algal blooms in both freshwater and marine environments (Seitzinger and Phillips, 2017). Excess N fertilizer can also cause problems for crop production by 347 348 promoting weeds. Weeds increase as nutrient availability increases (MacLaren et al., 2019), and so a high dose of N applied early in the season when crops are too small to capture it can be 349 350 expected to promote the establishment of weeds that will become competitive with crops later in 351 the season as resource availability diminishes.

352

353 *Effects of N fertilizer on soil bacterial community composition and function*

The results of this study suggest that applying more N fertilizer alters the soil-bacteria community 354 composition, which can affect soil carbon and N cycling (Buchkowski et al. 2015). Increased N 355 356 appeared to shift the community toward a group of bacteria that are less reliant on N-containing carbon sources (amines and amino acids; Figure 4). Fierer et al., (2012) also observed that the 357 bacterial communities become less reliant on organic forms of N as N fertilization increased, 358 suggesting a shift from oligotrophic to copiotrophic communities. Bacteria typically require a 25:1 359 carbon:nitrogen ratio, and so where sufficient mineral N is available to meet this need, bacteria 360 361 can decompose carbon-rich organic molecules more readily. In contrast, if mineral N availability is low, then bacteria are limited in the amount of organic matter they can decompose by the need 362 to acquire organic N. Such conditions also promote N release through mineralization (Mengel, 363 1996). 364

365 Abundant N, therefore, could increase the capacity of the bacterial community to 366 metabolize soil carbon and release it into the atmosphere. This outcome would not be beneficial

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367 to either farmers (loss of soil organic matter) or the environment (carbon emissions) (Hasselquist et al., 2012). It is not yet clear whether N fertilizer universally increases carbon emissions, as other 368 studies have observed that the addition of N in some instances negatively affects soil respiration, 369 370 leading to an overall increase in carbon sequestration (Janssens et al., 2010). N fertilizer can reduce both microbial biomass as well as activity, particularly in bacteria (Demoling et al., 2008), and 371 mainly in the presence of recalcitrant organic matter, while positive effects are observed when N 372 373 is added to easily degradable organic material (Fog, 1988). This study measured neither in situ 374 bacterial biomass nor activity and so further research will be necessary to determine if high N rates 375 do release carbon from South African soils.

In cases where N fertilizers do reduce carbon emissions by suppressing bacterial activity, then a negative effect on other functions can also be expected. This includes N mineralization, an important pathway by which N is released from soil organic matter and made available to plants. High N levels tend to reduce mineralization in general, whether or not they suppress bacterial activity, as microbes tend to take up more mineral N when it is in high quantities, thus delaying the mineralization of N from organic sources (Fog, 1988; Zhou et al., 2012).

382 Overall, a bacterial community that is more adapted to decomposing N-containing carbon sources in the presence of low mineral N would therefore be expected to release more N through 383 mineralization per unit of carbon respired. This quality could promote sustainability through 384 maximising nitrogen availability to crops via mineralization, while minimising carbon lost from 385 386 soil to the atmosphere. Our study suggests that such communities can be promoted by restricting the amount of N fertilizer. If the amount of N supplied to crops through mineralization rather than 387 fertilization can be increased, this could also reduce the large carbon footprint associated with 388 389 synthetic N fertilizers (Lal, 2004).

390 Despite the observed shift in carbon source utilization in this study, there was no effect of 391 fertilizer amount or distribution on either N fixation potential (number of nifH gene copies) or 392 bacterial community diversity. This contrasts with other studies, which have shown that N 393 fertilization can reduce the abundance of rhizobia (Ledgard, 2001) and free-living diazotrophs (Orr et al., 2011; Compton et al., 2004). N fertilization has also been observed to overall community 394 395 diversity (Ramirez et al., 2010, Coolon et al., 2013; Wang et al., 2015) and activity (Kennedy et al., 2004, Demoling et al., 2008). This negative effect of N on microbial activity is mainly found 396 in the presence of recalcitrant organic matter, while a positive effect is observed when N is added 397 to easily degradable organic material (Fog, 1988). That N fertilizer tends to raise soil pH is at least 398 399 partially responsible for such trends (Kennedy et al., 2004; Lauber et al., 2009; Wang et al., 2015). 400 Such effects were not observed in this study, possibly as a result of the history of the trial sites. Perhaps local diversity and N-fixing bacteria were already depleted after decades of intensive 401 402 cropping. Bacterial diversity and community composition may also depend on other soil qualities 403 (Williams et al., 2013) and it is not known whether these were limiting at any or all of the sites in 404 the present study. The effect of adding N to a microbial system remains difficult to model and explain (Hasselquist et al., 2012; Janssens et al., 2010), with different studies often producing 405 406 apparently contradictory results (e.g. Williams et al., 2013; van der Bom et al., 2018). This emphasizes the importance of continuing studies to elucidate the complex relationships between 407 408 farm management, microbial communities, and carbon and nutrient cycling. However, increased awareness of the effect of inorganic N on microbial biomass and activity will inform better 409 410 management practices.

411

412 Conclusion

This study suggests that applying less N fertilizer more often over a cropping season is optimal for both crop production and environmental protection, and may help to sustain the capacity of the soil bacterial community to contribute to both. Applying 150 kg ha⁻¹ of N fertilizer compared to 60 kg ha⁻¹ in this study did not increase canola yield, and splitting the fertilizer into three applications throughout the season reduced N losses, which may have been due to leaching and/or
bacterial immobilisation. This increased the soil mineral N available later in the season, where it
may contribute to higher yield quality. Furthermore, applying a large amount of N fertilizer
appeared to shift the community toward taxonomic groups that are more prone to immobilize soil
N and release soil carbon.

Applying a high level of N fertilizer is thus a 'lose-lose-lose' situation for farmers, their soil ecosystems, and the environment. Fortunately, farmers in the Western Cape tend to apply 70-90 kg ha⁻¹ and some do split it over two or three applications in the season. This study confirms the wisdom of such practices: canola does not use additional N, and excess N may affect the functioning of agricultural soil and threaten natural ecosystems. Ongoing trials in the region will assess the response of canola to a greater range of fertilization strategies (du Toit, 2018) to allow the optimal amount and distribution of N fertilizer for canola to be refined.

429

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598 FIGURES

Figure 1: Soil mineral nitrogen (N) and canola biomass and leaf area index (LAI) in each 599 treatment at 60, 90 and 120 days after emergence (DAE). LAI was not measured at 120 DAE. 600 601 Labels along the x-axis refer to the different treatments: 'Cont' = Control, and other labels indicate the total N rate in kg ha⁻¹ (60 or 150) and the number of applications in which this was 602 applied (1, 2 or 3). Soil samples were taken prior to fertilization. Error bars indicate the standard 603 604 error of the mean. Lowercase letters above the bars indicate pairwise differences significant at 605 P=0.05, based on contrasts estimated from the ANOVAs in Table 4. Treatments that differ do not share the same letter. Shading denotes the control treatment (pale), the 60 kg N ha⁻¹ treatments 606 (medium) and the 120 kg N ha⁻¹ treatments (dark). Note the log scales on the y-axes. 607

608

Figure 2: Canola yield in response to N fertilizer rates and timing. Labels along the x-axis refer to the different treatments: 'Cont' = Control, and other labels indicate the total N rate in kg ha⁻¹ (60 or 150) and the number of applications in which this was applied (1, 2 or 3). Lowercase letters indicate pairwise differences significant at P=0.05, based on contrasts estimated from the ANOVAs in Table 5. Treatments that differ do not share the same letter. Shading denotes the control treatment (pale), the 60 kg N ha⁻¹ treatments (medium) and the 120 kg N ha⁻¹ treatments (dark).

616

Figure 3: The NMS ordination of the OTUs of bacterial DNA extracted from soil in each
treatment at the end of the season. The amount of N applied to each plot is denoted by shading,
while the number of N applications is not indicated as the community did not differ in relation to
application number.

622 Figure 4: Axes 1 and 2 (a) and 1 and 3 (b) of the NMS ordination of carbon source utilization by bacterial communities from soil in each treatment at the end of the season. The amount of N 623 applied to each plot is denoted by shading (see legend), while the number of N applications nor 624 625 farm are indicated as these did not affect carbon utilization. The blue letters indicate the mean centroid of each N amount treatment (H = high; 150 kg N ha⁻¹, L = low; 60 kg N ha⁻¹ and N = 626 627 none; control), and the blue circle indicates the standard error of that mean. The boxed labels refer to the carbon sources most used by sites in that part of the ordination. Labels further away from 628 the centre of the plot indicate a greater change in the use of that carbon source in that direction. 629 a. acids = amino acids; c. acids = carboxylic acids; carbs = carbohydrates; phos.cpds = 630 phosphorylated compounds 631

- **Table 1:** Soil chemical and physical characteristics of the research sites, Langgewens Research
- Farm, Altona and Roodebloem. Samples were taken prior to onset of the trial and were taken to a
- 635 depth of 150 mm.

	Langgewens	Altona	Roodebloem
pH(KCl)	5.8	6.3	5.8
Calcium (mg kg ⁻¹)§	944	1572	1250
Magnesium (mg kg ⁻¹)§	160	191	204
Potassium (mg kg ⁻¹)§	230	209	463
Phosphorus (mg kg ⁻¹) [§]	111	73	116
Sulphur (mg kg ⁻¹) ¶	32	25	8.8
Mineral Nitrogen (mg kg ⁻¹)*	880	3040	2080
Organic Carbon (%)	0.94	1.11	1.58
Textural class	Sandy loam	Sandy loam	Sandy loam
Sand (%)	63	57	71
Slit (%)	30	18	14
Clay (%)	7	25	15

636 [§]Citric acid extraction (Non-affiliated Soil Analysis Work Committee, 1990)

637 [¶]Calcium phosphate extraction (Non-affiliated Soil Analysis Work Committee, 1990)

^{*}Indophenol-blue and salicylic acid methods (Cataldo et al., 1975; Keeney et al., 1982)

640 **Table 2:** Nitrogen fertilization rates and distribution for canola production at planting, 30 days

641 after emergence (DAE), 60 DAE and 90 DAE. Treatment 0 is the control, which received no

Treatment		Amount of N applied at each time point (kg ha ⁻¹)						
Rate (kg N ha ⁻¹)*	Number of applications**	At planting	30 DAE	60 DAE	90 DAE			
0	0	0	0	0	0			
60	1	20	40	0	0			
60	2	20	20	20	0			
60	3	20	13.3	13.3	13.3			
150	1	20	130	0	0			
150	2	20	65	65	0			
150	3	20	43.3	43.3	43.3			

642 nitrogen (N) fertilization throughout the season.

⁶⁴³ * Refers to the total amount applied over the whole season

**Refers to the number of applications after planting, as all treatments (except the control)
received 20 kg ha⁻¹ N at planting, following common practice in the region.

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648

649 **Table 3:** Summary of variables measured in this trial and the time point(s) at which they were

650 measured. DAE = Days after emergence

Response variables	60 DAE	90 DAE	Physiological maturity/harvest (approximately 150 DAE)
Soil N and	Soil mineral N	Soil mineral N	Soil mineral N
canola growth	Biomass	Biomass	Biomass production
measurements	LAI	LAI	Yield
			Harvest index
			Nitrogen (crude protein)
Bacterial			OTUs (ARISA)
community			Carbon utilization
measurements			nifH gene copies

651

653 Table 4: ANOVA F-statistics and P-values for the fixed effects in the mixed models of soil mineral 654 N and canola leaf area index and biomass, and standard deviations for each random effect and the 655 residual. Significant P-values <0.05 for fixed effects are emphasized in bold. Where cells are 656 blank, data was not collected for that response at that time point. DAE=days after emergence.

Time point	Effect type	Log soil	min	eral N	Log leaf	f area	index	Log	biom	ass
	Fixed effects	F statistic	df	P value	F statistic	df	P value	F statistic	df	P value
	Control vs treated	1.888	66	0.174	4.518	65	0.037	4.064	65	0.048
60 DAE	N amount	0.488	66	0.487	10.344	65	0.002	11.721	65	0.001
	N distribution	2.561	66	0.085	1.654	65	0.199	1.89	65	0.159
	Interaction*	0.847	66	0.433	4.964	65	0.01	6.34	65	0.003
	Random effects	Standard d	eviat	ion	Standard d	eviati	on	Standard d	eviati	on
	Replicate	0.078			0.000			0.000		
	Farm	0.285			0.456			0.706		
	Residual	0.356			0.268			0.442		
	Fixed effects	F statistic	df	P value	F statistic	df	P value	F statistic	df	P value
	Control vs treated	26.101	66	<0.001	7.391	65	0.009	3.954	65	0.051
90 DAE	N amount	3.732	66	0.058	5.772	65	0.019	10.213	65	0.002
	N distribution	6.884	66	0.002	1.626	65	0.205	1.481	65	0.235
	Interaction*	0.535	66	0.588	1.253	65	0.293	3.204	65	0.047
	Random effects	Standard d	eviat	ion	Standard d	eviati	on	Standard d	eviati	on
	Replicate	0.046			0.104			0.000		
	Farm	0.412			0.249			0.202		
	Residual	0.359			0.400			0.551		
	Fixed effects	F statistic	df	P value	F statistic	df	P value	F statistic	df	P value
Physio-	Control vs treated	37.97	66	<0.001				2.425	65	0.124
logical	N amount	8.541	66	0.005				1.885	65	0.175
maturity	N distribution	14.867	66	<0.001				0.338	65	0.714
	Interaction*	4.713	66	0.012				0.977	65	0.382
	Random effects	Standard d	eviat	ion	Standard d	eviati	on	Standard d	eviati	on
	Replicate	0.155						0.059		
	Farm	0.469						0.262		
	Residual	0.334						0.366		

657 _____*The N amount x N distribution interaction.

Table 5: ANOVA F-statistics and P-values for the fixed effects in the mixed models of canolayield, harvest index and tissue N content, and standard deviations for each random effect and the

660 residual. Significant P-values (< 0.05) for fixed effects are emphasized in bold. Where cells are

blank, data was not collected for that response at that time point. DAE=days after emergence.

		Yield		Harv	est ind	lex	Tissue	N co	ntent
Fixed effects	F statistic	df	P value	F statistic	df	P value	F statistic	df	P value
Control vs treated	20.575	65	<0.001	0.017	65	0.896	1.251	64	0.267
	1.058	65	0.308	1.634	65	0.206	1.537	64	0.22
N amount									
	0.013	65	0.987	0.557	65	0.576	0.243	64	0.785
N distribution									
	0.032	65	0.969	1.315	65	0.275	0.138	64	0.871
Interaction*									
Random									
effects	Standard de	viation		Standard dev	viation		Standard de	viatio	n
Replicate	0.000			0.000			0.064		
Farm	0.481			0.094			0.187		
Residual	0.956			0.085			0.247		

662

*The N amount x N distribution interaction.

663

664 Table 6: ANOVA F-statistics and P-values for the amount and distribution of nitrogen (N) in the 665 mixed models of bacterial richness, Shannon diversity and copies of the nifH gene in the soil when 666 the canola reached physiological maturity. Significant P-values <0.05 are emphasized in bold.</p>

	OTU Shannon diversity			ΟΤυ	rich	iness	nifH		
Fixed effects	F statistic	df	P value	F statistic	df	P value	F statistic	df	P value
Control vs									
treated	1.598	58	0.211	1.513	58	0.224	0.002	75	0.969
N amount	0.162	58	0.689	0.224	58	0.638	0.714	75	0.401
N distribution	1.459	58	0.241	2.579	58	0.084	2.324	75	0.106
Interaction*	1.748	58	0.183	3.22	58	0.047	1.513	75	0.228
Random									
effects	Standard de	eviat	ion	Standard of	levia	ation	Standard d	evia	tion
Replicate	0.000			0.00			0.0		
Farm	0.796			12.35			629.4		
Residual	1.091			14.98			1059.0		

*The N amount x N distribution interaction.

669 **Table 7:** PERMANOVA F-statistics and P-values for the dissimilarity matrices of bacterial DNA

and carbon source functional composition. The PERMANOVAs were calculated on the same

- 671 Bray-Curtis dissimilarity matrices used for the NMDS ordinations (Figures 4 and 5). Significant
- 672 P-values < 0.05 are emphasized in bold.

	Bacterial DNA		Community fu	nction		
	PERMANOVA F	P Value	PERMANOV A F	P Value		
Control vs treated	1.167	0.079	1.8	0.124		
N amount	1.552	0.007	2.564	0.051		
N distribution	0.875	0.316	1.248	0.272		
Interaction*	1.185	0.031	1.468	0.169		

673

*The N amount x N distribution interaction.

















3 Figure 4

2