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Growth of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* for biodiesel production and carbon dioxide capture

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Growth of
***Chlorella vulgaris* and *Chlamydomonas reinhardtii* for**
biodiesel production and carbon dioxide capture

by

Mariana T. Farcas

Thesis submitted to the
Davis College of Agriculture, Forestry, and Consumer Sciences
at West Virginia University
in partial fulfillment of the requirements
for the degree of

Master of Science
in
Applied and Environmental Microbiology

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2012

Keywords: growth, *Chlorella vulgaris*, *Chlamydomonas reinhardtii*, biodiesel, carbon dioxide capture, immobilized algae.

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ABSTRACT

Growth of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* for biodiesel production and carbon dioxide capture

by Mariana T. Farcas

The growth of two strains of green microalgae, *Chlorella vulgaris* (UTEX 2714) and *Chlamydomonas reinhardtii* (UTEX 90) was tested in three types of media; Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC buffered with either glycylglycine or Tris-base). Also, initial medium pH is ranging from 4 to 10, light intensity ranging from 100 to 600 $\mu\text{mol photons/m}^2\text{s}$, and CO_2 concentrations ranging from 0.038% (ambient) to 12%, were tested. WC medium at pH 8 buffered with glycylglycine sustained the highest yield and best buffering capacity for growth of both *C. vulgaris* and *C. reinhardtii*. A light intensity of 200 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ provided for both good growth and electron transport rate (ETR). Both *C. vulgaris* and *C. reinhardtii* produced highest final yields when grown with 6% CO_2 . Also, lipid content increased with increasing CO_2 concentration. Myristoleic acid (C14:1), palmitic acid (C16:0), stearic acid (C18:0), oleic acid (C18:1n9), linoleic acid (C18:2), and docosahexaenoic acid (DHA) were found in higher content when *C. vulgaris* was grown on 12% CO_2 , while the content of palmitoleic acid (C16:1), elaidic acid (C18:1t9), vaccenic acid (C18:1n7) were similar among all CO_2 concentration tested. CO_2 capture was explored using two approaches: consumption of known quantities of CO_2 in sealed serum bottles, and consumption of CO_2 flowing through immobilized algal beads. In both cases, fixation rate decreased with increasing CO_2 concentration. CO_2 consumption generally decreased over the five day experiment. The rate observed using immobilized algae was 20% of the maximum obtained in liquid culture, indicating the need to future optimize this novel method for CO_2 capture.

DEDICATION

I dedicate this thesis to my children and my husband. Without their love, encouragement, guidance, and support I never would have achieved this goal.

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GENERAL INTRODUCTION

As a result of the industrial revolution in the 20th century, an increased volume of CO₂ has been released into earth's atmosphere. The upper safety limit for atmospheric CO₂ is 350 parts per million (ppm) and since 1988 this limit has been exceeded (National Oceanic and Atmospheric Administration). CO₂ concentration in the atmosphere is still increasing and the outcome of this increase has already had a profound effect on the global environment. It is widely accepted that global warming is impacted by greenhouse gas emissions from anthropogenic activities (Intergovernmental Panel on Climate Change, 2005). Thus actions are being taken to mitigate greenhouse gas emissions from anthropogenic activities (Gough, 2008). There are three types of mitigations strategies: (1) chemical reaction-based approaches: washing with alkaline solutions, multiwalled carbon nanotubes, and amine coating activated carbon; (2) direct injection underground or to the ocean; and (3) biological CO₂ mitigation, with CO₂ being biologically converted to organic matter (Ho *et al.*, 2011). Biological CO₂ fixation is accomplished by all terrestrial plants and photosynthetic microorganisms through the process of photosynthesis. However, plants are expected to account for only a 3–6% reduction in global CO₂ emissions (Skjanes *et al.*, 2007). Costa *et al.* (2000) estimated that due to faster growth of microalgae and cyanobacteria, CO₂-fixation efficiency should be 10–50 times greater. Microalgal biomass also could be used as a feed stock for a variety of biofuels, medications, cosmetics, and nutritious foods, representing additional benefits from the microalgal CO₂ reduction process (de Morais and Costa, 2007).

In 2010, the U.S. Department of Energy funded twelve large projects (more than \$107 M total) for “Innovative Concepts for Beneficial Reuse of Carbon Dioxide”. Six of these projects are systems that use microalgae to capture CO₂ from power plant flue gas and convert it (via sunlight, water, and nutrients) into natural oils that are readily processed into liquid transportation fuels such as biodiesel. Photosynthetic green algae are good candidates to capture excessive amounts of atmospheric CO₂, since these organisms are capable of fixing CO₂ to produce energy and chemical compounds (fatty acids) upon exposure to sunlight.

Under optimal conditions of growth, fatty acids of microalgae constitute about 5–20%

of dry cell weight (DCW), but under unfavorable environmental or stress conditions for growth, algae alter their lipid biosynthetic pathways towards accumulation of neutral lipids (20–50% DCW), in the form of triacylglycerol (TAGs) (Guschina and Harwood, 2006; Hu *et al.*, 2008; Thompson, 1996). Based upon the photosynthetic efficiency and growth potential of algae, theoretical calculations indicate that annual oil production can be 100-fold greater than that of soybeans. However, few systems of algae-based biofuel production have progressed beyond the small laboratory or field testing stages (Hu *et al.*, 2006).

The overall objective of the current project is to build an effective system to grow the green algae *Chlorella vulgaris* and *Chlamydomonas reinhardtii* for CO₂ capture and the production of biodiesel. Experiments were designed to compare photoautotrophic growth rates and biomass production by *Chlorella vulgaris* and *Chlamydomonas reinhardtii* with varying conditions of CO₂, light, and nutrient availability. This information was utilized to design two growth systems to investigate capture of CO₂ by *Chlorella vulgaris* and *Chlamydomonas reinhardtii* using: 1) fed-batch growth; and 2) a flow-through system based on immobilized algae in alginate beads.

REFERENCES

Costa JAV, Linde GA, Atala DIP, Mibielli GM, Krüger RT. 2000. Modelling of growth conditions for cyanobacterium *Spirulina platensis* in microcosms. *World J Microbiol Biotechnol.* **16**: 15-18.

Gough C. 2008. State of the art in carbon dioxide capture and storage in the UK: An experts' review. *International Journal of Greenhouse Gas Control.* **2**: 155–168.

Guschina IA, Harwood JL. 2006. Lipids and lipid metabolism in eukaryotic algae. *Prog. Lipid Res.* **45**: 160–186.

Ho S-H, Chen C-Y, Lee D-J, Chang J-S. 2011. Perspectives on microalgal CO₂-emission mitigation systems — A review. *Biotechnology Advances.* **29**: 189-198.

Hu Q, Zhang CW, Sommerfeld M. 2006. Biodiesel from Algae: Lessons Learned Over the Past 60 Years and Future Perspectives. Juneau, Alaska: Annual Meeting of the Phycological Society of America, July 7–12, pp. 40–41 (Abstract).
<http://www3.interscience.wiley.com/journal/118594641/abstract>

Hu Q, Sommerfeld M, Jarvis E, Ghirardi M, Posewitz M, Seibert M, Darzins Al. 2008. Microalgal triacylglycerols as feedstocks for biofuel production: perspectives and advances. *Plant Journal.* **54**: 621-639.

de Morais MG, Costa JAV. 2007. Biofixation of carbon dioxide by *Spirulina* sp. and *Scenedesmus obliquus* cultivated in a three-stage serial tubular photobioreactor. *J. Biotechnology.* **129**: 439–445.

Skjanes K, Lindblad P, Muller J. 2007. BioCO₂ – A multidisciplinary, biological approach using solar energy to capture CO₂ while producing H₂ and high value product. *Biomol Eng.* **24**:405-413.

Thompson, G.A. 1996. Lipids and membrane function in green algae. *Biochim. Biophys. Acta.* **1302**: 17–45.

U.S. Department of Energy. 2010. Innovative concepts for beneficial reuse of carbon dioxide. http://fossil.energy.gov/recovery/projects/beneficial_reuse.html

CHAPTER 1

Growth of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* with varying growth media, pH, light, and CO₂ concentration

1.1. INTRODUCTION

Unicellular microalgae capture light energy for CO₂ fixation and biomass production. A portion of this biomass accumulates as triacylglycerols, which can be harvested to produce biodiesel transportation fuels (Francisco *et al.*, 2010). In laboratory experiments, growth conditions for production of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* associated with fatty production were investigated.

Early attempts in the late 1800's to culture microalgae were reviewed by Allen and Nelson (1910) including basic media developed by Farmintzin, Beijerinck and Miquel. These works first described the significance of culture pH, iron, and vitamins for healthy algal growth and also the relationship to amounts of dissolved oxygen and carbonic acid, metal toxicity, and impurities in distilled water (Allen and Nelson, 1910). In the USA, the Stanford Research Institute was the first to attempt large-scale algal culture. They demonstrated the technical feasibility of *Chlorella* sp. production as an industrial raw material (Burlew, 1976). Spoehr and Milner (1949) first studied the effect of environmental conditions on chemical composition of *Chlorella* sp., and, in particular, the fatty acids profile of this alga. They found that nitrogen limitation increases lipid content.

Microalgae frequently are cultivated in batch culture. An algal inoculum is added to a specific growth medium containing nutrients necessary for growth until stationary phase is reached and the total culture is harvested (Richmond, 2004). Optimal batch growth conditions for microalgal cultures are strain specific, and final biomass production is influenced by many factors. The most important abiotic parameters regulating algal growth are: light cycle and intensity, temperature, nutrient quantity and quality, pH, carbon dioxide, salinity, and biotic factors like cell fragility and cell density. These factors affect photosynthesis, influence the pattern, pathway and activity of cellular metabolism and as a result dynamic cell composition (Anderson, 2005). Mechanical factors affecting growth include mixing, gas bubble size and distribution, and mass transfer (Schenk, 2008).

Growth of microalgae in culture media requires a suitable composition of essential macronutrients and micronutrient (Anderson, 2005). For autotrophic growth, supplemental CO₂/HCO³⁻ often is necessary to supply the C-requirements of high yield autotrophic algal production. According to Redfield ratio, the stoichiometric ratio of the phytoplankton is C: N: P=106:16:1, thus, most media are nitrogen-rich relative to carbon, and carbon can become limiting (Riebesell *et al.*, 1993).

The carbonate-bicarbonate system is the main buffer to control and uphold appropriate pH levels optimum for algal culture (Richmond, 2004). If culture pH increases quickly to 9 or higher, this may be a signal that carbon may be limiting. To prevent a reduced growth rate or cell yield, bubbling with CO₂ or adding more bicarbonate in late exponential phase often is used. Previous studies have shown that the CO₂ supply needs to be optimized, because higher levels can cause growth inhibition and low levels can limit growth (Rados *et al.*, 1975). These maximum (inhibition) and minimum (limitation) concentrations vary among species. The most studied technique of supplying CO₂ is by injection of CO₂ mixed in the airflow (Chini *et al.*, 1996) or directly in the culture medium (Molina-Grima *et al.*, 1994).

Algal productivity requires the essential nutrients nitrogen and phosphorus (Schenk, 2008). The usual nitrogen sources in algal media are ammonium, nitrate, or urea. Special attention is required for ammonium as a sole source of nitrogen since the culture pH can drop sharply during active growth due to the release of H⁺ ions, especially in dense cultures at high temperature. When nitrate is used as the sole nitrogen source, an increase in pH occurs (Richmond, 2004). Urea can also serve as a good nitrogen source once it is hydrolyzed, by the action of either the enzyme urease, or the enzyme urea amidolyase (Hodson, 1975). The major source of phosphorus in algal growth media is inorganic phosphates.

Other than C, N, and P, significant nutrients for algal nutrition are K, Ca, Mg, Cl, Mn, S, and Na in ionic form. They are constituents of chlorophyll (Mg) and are involved in photosynthesis (Cl, Mn), and cell metabolism (K, Na, Ca, S). Trace elements like Zn, Cu, Mo, and Mn are important nutrients involved in redox reactions (Taiz and Zeiger, 2002).

Some algae also require vitamins (Croft, 2006). The most common are vitamin B₁₂ (cyanocobalamin), thiamine, and biotin. Usually, only few algae need all three vitamins (Provasoli and Carlucci, 1974).

Buffers are important components of algal media. The most common pH buffers used to prevent or reduce metal precipitation and to maintain constant pH values are: TRIS (2-amino-2-(hydroxymethyl)-1-3-propanediol) and glycylglycine (2-[(2-aminoacetyl) amino] acetic acid) (McLachlan, 1973). Comparing TRIS and glycylglycine with other buffers like MOPS (3-N-morpholino propanesulfonic acid), HEPES (N-[2-hydroxyethyl] piperazine-N'-[2-ethanesulphonic acid]), and TAPS ([2-hydroxy-1, 1 bis (hydroxymethyl) ethyl] amino)-1-propanesulfonic acid]), Loeblich (1975) noted that TRIS and TAPS provided maximal growth with minimal pH change. TRIS also can serve as carbon source for contaminant bacteria, interfere with the analysis of dissolved organic nitrogen and ammonium, and toxic to some algal species (Fabregas *et al.*, 1993). TRIS can have harmful effects on photosynthesis by inhibiting the mechanisms of HCO_3^- transport across the plasma membrane (Axelsson *et al.* 2000; Hellblom *et al.*, 2001) or O_2 evolution in Photosystem II (Sofrova *et al.* 1978; Rickert *et al.* 1991).

Light energy is captured during photosynthesis for autotrophic CO_2 fixation. The photosynthetic process occurs in two phases, a photochemical phase that is light dependent and a biochemical dark phase that is light independent. ATP and NADPH produced during the light dependent phase are essential in the dark phase to synthesize molecules essential for growth (Laval and Mazliak, 1995). The optimal light intensity varies among different species, and the growth rate is lower at light intensities values below this value (Ojala, 1993). At high light intensities growth can be limited by the phenomenon of photoinhibition (Bouterfas *et al.*, 2006). Therefore, the intensity, spectral quality and photoperiod of light must be considered because all can impact the circadian rhythm of photosynthesis, respiration (Piquemal, 1990), cellular division (Hobson *et al.*, 1979), growth rate (Redalje and Laws, 1983), and enzymatic activities (Hobson *et al.*, 1979). Light can be natural or provided by fluorescent tubes emitting either in the blue or the red light spectrum able to support photosynthesis. Care must be taken with artificial lights to prevent overheating of cultures. Algae exhibit a growth rate proportional to the duration of the effective light period (Foy and Gibson, 1993), and also a light/dark cycle is needed for a maximal growth.

Mechanical factors like mixing and aeration can influence optimal algal growth. In batch cultures, mixing is important for preventing sedimentation of algal cells and to ensure all algal cells are equally exposed to the light and nutrients (Molina-Grima *et al.*, 1999). Air bubbles can

damage microalgae (Eriksen, 2008) due to shear sensitivity caused by eddies in the growth medium. To avoid cell adhesion to gas bubbles and reduce shear damage (cell death), cultures can be supplemented with non-ionic surfactants (Ma *et al.*, 2004).

Biofuels are defined as solid, liquid, or gaseous fuels that are produced from biorenewable feedstock (Demirbas, 2009). As fossil hydrocarbons become limited and costly, conversion of algal biomass into liquid biofuels is becoming more cost-effective and attractive (Demirbas, 2011). The use of domestically produced biofuels can reduce political and economic vulnerability, reduce greenhouse gas emissions, and invigorate the economy by creating a new energy sector (Balat, 2010). For example, algae could use CO₂ from power plant emissions for conversion to biomass and into biofuels (Aksoy *et. al.*, 2011). Studies showed the potential of microalgae to reduce CO₂ levels from industrial waste gas with an increase in biomass (Yun, 1997). There are advantages and disadvantages of biofuel production by microalgae. One advantage is the ability of algae to utilize less land resources and water than conventional agricultural crops. Another advantage is the high efficiency of CO₂ mitigation. The main problems that impede the commercial implementation of algae for biofuel production are the current low yield and high cost of biomass harvest, high capital costs to establish commercial facilities, and the increased care required by a microalgal farming facility compared to a conventional agricultural farm (Demirbas, 2011).

Previous studies (Converti *et al.*, 2009; Rodolfi *et al.*, 2009) have shown that biochemical composition of microalgae and cell lipid content can vary as a result of changes in environmental manipulations (growth conditions, nutrient concentration, and nutrient availability). The purpose of the preliminary work was to study the growth of two strains of green microalgae, *Chlorella vulgaris* (UTEX 2714) and *Chlamydomonas reinhardtii* (UTEX 90). Three types of media Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC), a pH range from 4 to 10, light intensity from 100 to 600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$, and CO₂ concentration of 0.038% (ambient), 3%, 6%, and 12%, were tested.

1.2.MATERIALS AND METHODS

1.2.1. Growth media and initial pH

Two strains of green microalgae *Chlorella vulgaris* (UTEX 2714) and *Chlamydomonas reinhardtii* (UTEX 90) were purchased from UTEX (The Culture Collection of Algae at the University of Texas at Austin). Both were maintained in under 12:12 light:dark cycle (200 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$) at $25\pm 2^\circ\text{C}$. Ambient air was bubbled through the vessels at $\sim 50 \text{ mL min}^{-1}$. The growth of both algae was compared in three media: Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes medium (WC).

Table 1 describes the chemical composition of Tris-Acetate-Phosphate (TAP) medium (Gorman and Levine, 1965). For the trace elements solution, EDTA was dissolved in boiling water, and the FeSO_4 was prepared last to avoid oxidation. All solutions were mixed, and EDTA was added last. Initially the mixture was green; however, after 1-2 weeks the solution eventually turned purple and could be used. Bushnell Haas Broth was purchased as a dehydrated medium (Difco, MD). The formula per liter is: MgSO_4 0.2g, CaCl_2 0.02g, K_2HPO_4 1g, KH_2PO_4 1g, NH_4NO_3 1g, and FeCl_3 0.05g. The medium was prepared by dissolving 3.27g in 1L distilled water, followed by sterilization at $121\text{-}124^\circ\text{C}$ for 20 minutes. The third medium used was Wright's Cryptophytes (WC) medium (Guillard and Lorenzen, 1972). Table 2 describes the chemical composition of WC medium. Either glycylglycine (WC-Gly) or Tris base (WC-Tris) were used as buffers. Table 3 contains the formula for the trace metal solution used in WC medium. All components were added to 950 mL of dH_2O , the final volume was adjusted to 1 liter with dH_2O , and was sterilized at $121\text{-}124^\circ\text{C}$ for 20 minutes. To prepare the vitamin solution (Table 4) used in WC medium, the thiamine was dissolved in 950 mL of dH_2O , 1 mL of the primary stock of biotin (vit. H) and cyanocobalamin (vit. B_{12}) were added and the final volume was adjusted to 1 liter with dH_2O . The final solution was filter sterilized and stored at 4°C . The initial pH of all three media (TAP, BHB, and WC) was adjusted to range from pH 4 to pH 10 using either 20% KOH or 1N HCl as needed.

Growth was compared using 24 vertical 200 ml glass test tubes attached to wood support. Six test tubes held 100 mL of TAP medium ranging from pH 4 to pH 10 (Fig. 1, Fig 2). Six more held BHB medium ranging from pH 4 to pH 10. Another twelve test tubes held WC medium ranging from pH 4 to pH 10, six of them using glycylglycine as buffer, and other six using Tris

base as buffer. The experiments were repeated twice. Two 460-watt metal GE grow lights (General Electric, Fairfield, CT) were suspended 1.5 m above the test tubes and supplied light ($200 \mu\text{mol photons m}^{-2}\text{s}^{-1}$) on a 12 hours light/12 hours dark schedule. Aluminum foil placed at a 45-degree angle at the base reflected light onto the test tubes. Un-inoculated controls were composed of a pH series of 4, 5, 6, 8, 9, and 10. The mouths of all tubes were closed using foam stoppers. Three mL of algal inocula ($\text{OD}_{750}=0.250$) (See Appendix 1) was injected into each cultures tubes using a syringe. Air (ambient CO_2 level = 0.038%) at 50 mL min^{-1} was bubbled through the testing tubes using a pipette to agitate the solution. Absorbance (OD_{750}) was measured every 24 hours during a 5-day period. The sampling was done by withdrawing three mL of algal suspension with a syringe, transfer to polymethacrilate cuvettes (Sigma-Aldrich, MO, USA), and determination the optical density (OD_{750}) using a Spectronic 20 Genesys spectrophotometer (Spectronic Instruments, NY). The pH was measured every 24 hours using a Corning 320 pH Meter (Corning Incorporated, NY).

1.2.2. Light intensity

A second type of photobioreactor was constructed to investigate the effect of light intensity on growth rate and photosynthetic efficiency. Three mL inocula of *C. vulgaris* and *C. reihardtii* ($\text{OD}_{750}=0.250$) (See Appendix 1) were injected in 100 mL WC medium (pH=8) in the same test tubes used in the previous experiments but placed on their sides on two shelves illuminated with a variable light source (460-watt metal GE grow lights source, P.L Light Systems). Light intensities were varied using a rheostat from $100\text{-}600 \mu\text{mol photons m}^{-2}\text{s}^{-1}$ (12 hours light/ 12 hours dark) (Fig. 3) and air ($\text{CO}_2=0.038\%$) was supplied at 50 mL min^{-1} . The experiments were repeated twice with three replicates for each algal species and light intensity. Control test tubes were run with air (ambient CO_2 level = 0.038%) at 50 mL/min in the same light condition but not inoculated. Absorbance was measured at 750 nm every 24 hours during a 5-day period. The sampling was done as described above.

Phytoplankton Analyzer (PHYTO-PAM, Heinz Walz GmbH, Germany) (Fig. 4) was used to measure the relative electron transport rate (ETR) as a function of PAR (Photosynthetically Active Radiation) on day two, three and four of the experiment. The measuring principle is based on selective amplification of a fluorescence signal which is

measured with the help of intense, but very short pulses of measuring light. Three mL of algal sample was withdrawn from the test tubes and dark adapted, by keeping the polystyrene fluorometer cuvettes (Sigma-Aldrich, MO, USA) in the optical unit for five minutes. The minimal fluorescence yield (F_o) was measured and immediately followed by the maximal fluorescence yield (F_m). When the actinic light provided by the halogen lamp was switched on, the minimum fluorescence in actinic light (F_t) was measured and followed by simultaneous measurement of the maximum fluorescence yield in actinic light (F_m').

The maximum efficiency of dark-adapted PSII (F_v/F_m) was calculated as:

$$F_v/F_m = F_m - F_o/F_m$$

The maximum efficiency of PSII in actinic light (Φ_{PSII}) was calculated as:

$$\Phi_{PSII} = (F_m' - F_t) / F_m'$$

Photosynthetic ETR ($\mu\text{mol m}^{-2} \text{s}^{-1}$) was calculated as:

$$\text{ETR} (\mu\text{mol electrons m}^{-2} \text{s}^{-1}) = \Phi_{PSII} \times \text{incident PAR} (\mu\text{mol photons m}^{-2} \text{s}^{-1}) \times 0.5 \times 0.84$$

where 0.5 is a correction factor based on the general assumption that half of the absorbed light energy was diverted to PSII, and 0.84 is absorptivity.

1.2.3. CO₂ concentration

To determine biomass production as a function of CO₂ concentration, a third type of photobioreactor was constructed. These photobioreactors consisted of six 160 mL glass serum bottles filled with 100 mL WC and sealed with solid stoppers (Fig. 5). Two needles were inserted through the stopper to bubble the gas through the medium and to provide an exit/sampling point. Laboratory air and CO₂ were mixed. CO₂ concentrations of ambient air (CO₂ level=0.038%), 3%, 6%, and 12% were established and verified by using a Carle AGC Series 100 gas chromatograph (Hach Co., Loveland, CO, USA) connected to a BD40 chart recorder (Kipp & Zonen, Inc., Bohemia, NY, USA). Light, held constant at $200 \mu\text{mol m}^{-2} \text{s}^{-1}$, was provided by two GE grow lights source (12 hours light/12 hours dark) (General Electric, Fairfield, CT, USA). Three mL inocula of *C. vulgaris* and *C. reinhardtii* (OD₇₅₀=0.250) (See Appendix 1) were injected in 100 mL WC medium (pH=8). The experiments were repeated twice with three replicates for each algal species and CO₂ concentration. Two un-inoculated bottles were incubated under the same condition and used as controls.

Absorbance and ETR measurements were done as described above (See 1.2.2.).

1.2.4. Lipid extraction

Lipid and fatty acid composition were compared as a result of bubbling different CO₂ concentrations (mixed with air) through the medium as the carbon source for photosynthesis. Two systems were used for lipid extraction: 8000M Mixer/Mill® (SPEX SamplePrep LLC, Metuchen, NJ, USA) and Soxtec™ 2055 System (Foss Analytical, Denmark).

1.2.4.1. Mixer/Mill system

The SPEX SamplePrep 8000 Mixer/Mill is an efficient, compact laboratory mill capable of pulverizing samples in the 10-gram range by shaking the containers back and forth at approximately 1080 cycles per minute (Fig. 6). It has been used for pulverizing rocks, minerals, sand, cement, slag, ceramics, and many others hard samples (8000M Mixer/Mill, Spex SamplePrep, Operating Manual).

Extractions were performed using freeze dried algal biomass. One gram of dried cells was transferred to a 25 ml stainless steel vial of a mixer/mill shake container. Twenty mL chloroform/methanol (2/1) was added to the container and the sample was shaken for 10 minutes (1425 rpm). Then the entire mixture was transferred to 30 mL glass tube and centrifuged at 5000 rpm for 5 min to recover the liquid phase. After cell disruption, the mixture was washed with 4 ml 5% NaCl solution and vortexed for few seconds. After this step, the mixture was centrifuged at low speed (2000 rpm) for 10 min to separate the two phases. The upper phase was removed using a Pasteur pipette. The supernatant was collected and the solvent was vaporized under a nitrogen stream. The total lipid content was then measured gravimetrically using an analytical balance with e=0.0001g (Denver Instrument XE-100, NY, USA).

1.2.4.2. Soxtec system

The Soxtec™ 2055 System (Foss Analytical, Denmark) represents a new patented version of the Soxhlet extraction technique, consisting of an Extraction Unit and a Control Unit (Fig 7). The system is designed for maximum user convenience through batch handling of six samples at a time.

The extraction was performed using freeze dried algal biomass. One gram of freeze dried sample was weighed into a thimble and a defatted cotton plug was place on top of the sample.

Next, the samples were inserted into the extraction unit using the thimble support holder. A mixture of 45 ml chloroform/methanol (2/1) was added to the glass containers and 4-5 glass beads to relieve hot spots. After the 3-step extraction procedure consisting of boiling (20 min at 120°C), rinsing (45 min), recovery (3 min), and drying (20 min), the extractable matter was collected in the cup. The total lipid content was measured gravimetrically as describe above.

1.2.5. Determinations of fat and fatty acids

The total fat and fatty acid content of the lipids were determined by the Agricultural Experimental Station, Chemical Laboratories, University of Missouri-Columbia. Samples were analyzed according to AOAC official method 996.06. Fat and fatty acids were extracted by the hydrolytic method and then methylated to fatty acid methyl esters (FAMES) using BF₃ in methanol, then quantified by capillary gas chromatography (GC).

1.3.RESULTS

1.3.1. Effect of growth medium and initial pH

Relative growth of *C. vulgaris* and *C. reinhardtii* was examined in four media: 1) Tris Acetate Phosphate (TAP); 2) Bushnell Haas Broth (BHB); 3) WC medium buffered with glycylglycine (WC-Gly); 4) WC medium buffered with Tris base (WC-Tris). The increase in OD₇₅₀ of both algae in each medium over a five-day period is reported in Figure 8 and 10. Initial experiments focused on TAP and BHB media. TAP is widely accepted as a laboratory growth medium for microalgae (Anderson, 2005; Harris, 1989), while BHB is a pre-formulated simple salts solution used to examine carbon source utilization by aquatic bacteria (Roy *et al.*, 2002). At day five, all three factors of variability: algae type, medium type, and initial pH had a significant effect on final OD₇₅₀ (See ANOVA table, Table 5). Overall, *C. vulgaris* displayed a greater growth response than *C. reinhardtii*. Between media types, TAP had a significantly higher final OD₇₅₀ compared with BHB. The pH that supported the greatest final growth was pH 7, followed by pH 8. Lowest final growth was observed at pH 4 and pH 10.

There was a significant interaction between algae type and medium type (Table 5), i.e. the effect of algae type on growth is dependent on the medium type. *C. vulgaris* had a higher overall OD₇₅₀ in TAP medium than BHB, while *C. reinhardtii* grew a better in BHB rather than TAP. The significant interaction between medium type and initial pH (Table 5) demonstrated

that TAP and BHB medium at pH 8 or 7 gave the highest growth. The lowest final OD₇₅₀ was observed at pH 4. The effect of algae type on growth was dependent on the initial pH (Table 5). *C. vulgaris* at pH 8 and pH 7 showed the highest response and *C. reinhardtii* indicated a better growth at pH 7. The lowest response was observed at pH 4 for both algae.

The three-way interaction between algae type, medium type, and initial pH also was significant (Table 5). A simple comparison of the final OD₇₅₀ obtained at day five for each alga in each medium is reported in Table 6. *C. vulgaris* grew optimally at pH 8 in TAP, while in BHB similar final growth was obtained at pH 6 and 7. In TAP *C. reinhardtii* grew equally well at pH 6 and 7, while in BHB had a similar growth response in the range of pH 6 to 8.

The growth responses described above in TAP and BHB media were complicated by the final pH measured in each culture vessel at day 5. Although un-inoculated controls maintained their initial pH readings throughout the incubation (Figure 12), inoculated samples generally exhibited a significant decrease of 0.5 to 1.0 pH units at the end of the growth experiment as a result of active algal growth (Figure 9 and 11, Table 7). In general, the observations of optimal growth at circumneutral pH in TAP and BHB are correct, but uncertainty of the actual medium pH at any given time compromises the interpretation of the statistical comparisons reported above. The data suggested the need for a more strongly buffered growth medium that would resist acidification resulting from algal growth. After an extensive literature review, a decision was made to test Wright's Cryptophyte Medium, buffered either with glycylglycine or Tris base. Superior growth was indeed observed in this medium (see below), however both WC-Gly and WC-Tris exhibited significant increases from initially acidic pH treatments, and significant decreases from initially basic conditions (range 0.5 to 3.1 pH units) over the 5 day incubation, again complicating an examination on the effect of pH on growth of these algae (Table 7 A, B).

Using WC medium, at day five all three factors of variability: algae type, medium type, and initial pH, had a significant effect on final OD₇₅₀ (See ANOVA table, Table 8). Overall growth of *C. vulgaris* in both WC media types was significantly higher than that of *C. reinhardtii*. Furthermore, WC-Gly medium supported better overall growth of both algae than did WC-Tris. An initial pH of 8 resulted in a significantly higher final OD₇₅₀ compared with all other pHs tested.

The effect of algae type on growth was dependent on the medium type (Table 8). Both algal species showed a significantly higher OD₇₅₀ in WC-Gly medium compared with WC-Tris.

Also, there was no significant difference between *C. vulgaris* growth in WC-Tris and *C. reinhardtii* growth in WC-Gly. The interaction between algae type and initial pH (Table 8), demonstrated that *C. vulgaris* grew better at pH 8, while *C. reinhardtii* had the highest growth at pH 7. The interaction between medium type and initial pH (Table 8) demonstrated that WC-Gly at pH 8 gave the best growth response, and the lowest growth was observed with WC-Tris at pH 4.

The three-way interaction between algae type, medium type and initial pH was significant (Table 8). According to the Tukey HSD test, *C. vulgaris* grown in WC-Gly medium at pH 8 had the greatest growth response. The best response for *C. reinhardtii* was obtained in WC-Gly medium at pH 8 (Table 6).

Analysis of all growth data obtained using all four media demonstrated that the three-way interaction between algae type, medium type and initial pH was significant (Table 8). *C. vulgaris* grown in WC-Gly medium at pH 8 had the highest final growth. The best growth response for *C. reinhardtii* also was obtained in WC-Gly medium at pH 8 (Table 6). Therefore further experiments reported below were performed by growing both algae in WC-Gly medium adjusted to an initial pH of 8.

1.3.2. Effect of light intensity

Relative growth and the relative electron transport rate of *C. vulgaris* and *C. reinhardtii* were examined at light intensities ranging from 100 to 600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$. The absorbance at 750 nm is reported in Figure 13. The relative electron transport rate (ETR) as a function of PAR (Photosynthetically Active Radiation) during day two, three and four is reported in Figure 14, 15, and 16. The growth and ETR were analyzed by multivariate analysis of variance (MANOVA), followed by univariate ANOVA using a fully factorial three-way model and planned contrast analyses. MANOVA indicated a significant effect on growth of algae type ($F_{(1,24)}=3.6844$, $p<0.0001$) and light intensity ($F_{(5,24)}=2.4308$, $p<0.0001$). Also, the interaction between algae type and light intensity is significant ($F_{(5,24)}=5.2018$, $p<0.0001$).

On day two, neither algae species showed a significant difference in growth response (Table 9), but on day three *C. vulgaris* had significantly higher growth comparing with *C. reinhardtii* (Table 10). This result also was observed on day four (Table 11) (Figure 13).

The overall effect of light intensity over time among both algae species, indicated that on day two the highest growth was observed at higher light intensities, however on day three and four a light intensity of 200 or 300 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ offered a better growth.

The interaction between algae type and light intensity was significant (Table 9, 10 and 11) and suggested that *C. vulgaris* outgrew *C. reinhardtii* at all the light intensity tested. Also, *C. reinhardtii* growth was not significantly different among all light intensity studied. In the case of *C. vulgaris*, growth at light intensity of 300 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ was significantly higher comparing to all other tested.

MANOVA showed a significant effect on ETR of algal type ($F_{(1,11)}=16.3654$, $p<0.0001$), and light intensity ($F_{(5,11)}=1.5951$, $p=0.0002$). Also, the interaction between algae type and light intensity was significant ($F_{(5,11)}=2.3229$, $p<0.0001$).

Overall, *C. reinhardtii* showed a significant higher ETR over time compared to *C. vulgaris* (Table 12, 13, and 14) (Fig 14, 15, and 16).

On day four, light intensities from 200 to 500 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ resulted in significantly higher ETR than 100 and 600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ (Table 14). ETR values obtained from *C. reinhardtii* at light intensities from 200 to 500 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ were significantly higher compared with results found at 100 and 600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$. ETR values for *C. vulgaris* were not significantly different among light intensities with the exception of 400 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$.

1.3.3. Effect of CO₂ concentration

Relative growth and the relative electron transport rate of *C. vulgaris* and *C. reinhardtii* were examined at different CO₂ concentrations: 0.038% CO₂ (ambient), 3% CO₂, 6% CO₂ and 12% CO₂. The absorbance at 750 nm (Figure 17) and the relative electron transport rate (ETR) as a function of PAR (Photosynthetically Active Radiation) during day two, three and four is reported in Figure 18, 19, and 20. Growth and ETR were analyzed by multivariate analysis of variance (MANOVA), followed by univariate ANOVA using fully factorial three-way models and planned contrast analyses. MANOVA indicated a significant effect on growth of algae type ($F_{(1,16)}=10.7660$, $p<0.0001$), and %CO₂ tested ($F_{(3,16)}=19.5426$, $p<0.0001$). However, there is not

significant interaction between algae type and %CO₂ over time ($F_{(3,16)}=0.5061$, $p=0.0805$, $\alpha=0.050$).

Overall, *C. vulgaris* showed significantly higher growth during entire experiment compared to *C. reinhardtii* (Table 15, 16, 17, and Fig. 17). Furthermore, the highest OD₇₅₀ was obtained when 6% CO₂ was bubbled into the bottles, followed by 3% CO₂, 12% CO₂, and 0.038% CO₂. On day four, the growth of *C. vulgaris* at 6% CO₂ was not significantly different from the response at 12% CO₂, and *C. reinhardtii* growth at 6% CO₂ was not significantly different from the response of *C. reinhardtii* at 3% CO₂. The lowest OD₇₅₀ was obtained during growth on ambient CO₂ for both algae (Table 17).

MANOVA indicated a significant effect on ETR of algae type ($F_{(1,16)}=19.1880$, $p<0.0001$), and different %CO₂ tested ($F_{(3,16)}=3.7597$, $p<0.0001$), and also a significant interaction between algae type and %CO₂ over time ($F_{(3,16)}=3.4448$, $p<0.0001$, $\alpha=0.050$). Overall, ETR among both algae at CO₂ concentration tested decreases from day two (ETR=69), day three (ETR=54) and to day four (ETR=35).

During the entire experiment, *C. reinhardtii* showed a significant higher ETR compared to *C. vulgaris* (consistent with the results determined in the light experiment) (Fig. 18, 19, 20 and Table 18, 19, and 20). The highest ETR was obtained when 0.038% CO₂ was bubbled into the bottles and then decreases with increasing CO₂ concentration (exception day four, when ETR was higher at 6% CO₂ but not significantly different from ambient and 3% CO₂. Determinations on day three and four indicated that 12%CO₂ negatively affected the ETR.

1.3.4. Total lipid content of *C. vulgaris* and *C. reinhardtii* grown at ambient, 6% and 12% CO₂

The total lipid contents for the microalgae cultured in this study ranged from 8.57% to 17.79% of the dry weight (Table 23). ANOVA (Table 21) showed that the algal type and CO₂ concentration had a significant effect on lipid content, but the extraction type was not significant. The total lipid content of *C. vulgaris* (14.9121 g lipid/ 100g algae biomass) was significantly higher compared to *C. reinhardtii* (12.5746 g lipid/ 100g algae biomass). Furthermore, a higher CO₂ concentration resulted in significantly higher lipid content. The total lipid content at 12% CO₂

was 16.0475 g lipid/ 100g algae biomass, at 6% CO₂ the lipid content was 13.8056 g lipid/ 100g algae biomass, and at ambient CO₂ the lipid content was 11.3769 g lipid/ 100g algae biomass.

1.3.5. Total fat content of *C. vulgaris* and *C. reinhardtii* grown at ambient, 6% and 12% CO₂

Analyzing the percentage of total fatty acids per 100mg biomass, ANOVA indicated that algal type and CO₂ concentration had a significant effect on the total fat content, but the extraction type was not significant (Table 22). The total fat content of *C. vulgaris* (10.61 g fat/ 100g algae biomass) was significantly higher compared to *C. reinhardtii* (8.67 g fat/ 100g algae biomass). Furthermore, the data indicate that higher CO₂ concentrations resulted in significantly higher fat content. The total fats at 12% CO₂ were 11.3839 g/ 100g algae biomass, at 6% CO₂ the total fatty acids was 9.8777 g/ 100g algae biomass, and at ambient CO₂ total fats were 7.6600 g fat/ 100g algae biomass. There was no effect of algal type on the fat content depending on the level of extraction type, and no effect of algae type on the fat content depending on the level of %CO₂.

1.3.6. Fatty acids profiles of *C. vulgaris* and *C. reinhardtii* grown at ambient, 6% and 12% CO₂

Generally, linoleic acid (C18:2) and palmitic acid (C16:0) were found in a significantly higher concentration, followed by linolenic acid (ω18:3), elaidic acid (C18:1t9) and oleic acid (C18:1n9) ($F_{(9,29)}=386.6372$, $p<0.0001$). Fig. 21 presents the major fatty acid composition of *C. vulgaris* at different CO₂ concentrations.

The higher CO₂ concentrations had a significant effect on each fatty acids of *C. vulgaris* ($F_{(2,29)}=11.1242$, $p<0.0001$) (Fig. 21). Myristoleic acid (C14:1), palmitic acid (C16:0), stearic acid (C18:0), oleic acid (C18:1n9), linoleic acid (C18:2, and DHA were found in higher content when *C. vulgaris* was grown on 12% CO₂, while the content of palmitoleic acid (C16:1), elaidic acid (C18:1t9), vaccenic acid (C18:1n7) were similar among all CO₂ concentration tested. On the contrary, linolenic acid (ω18:3) was greater when *C. vulgaris* was grown at ambient CO₂ rather than higher CO₂ concentrations.

For *C. reinhardtii*, linoleic acid (C18:2) and palmitic acid (C16:0) are found in a significant higher concentration, next oleic acid (C18:1n9) ($F_{(9,29)}=65.1665$, $p<0.0001$) Fig. 22 presents the

major fatty acid composition of *C. reinhardtii* at different CO₂ concentrations. Similar to *C. vulgaris*, the CO₂ concentrations have a significant effect on each fatty acids of *C. reinhardtii* ($F_{(2,29)}=24.0857$, $p<0.0001$) (Fig. 22). Cells grown at 12% CO₂ and 6% CO₂ will results with a higher content of fatty acids compared with cells grown at ambient CO₂. Palmitic acid (C16:0), stearic acid (C18:0), oleic acid (C18:1n9), linoleic acid (C18:2), and DHA is found in higher content when *C. reinhrdtii* is grown at 12% CO₂. The CO₂ concentration did not effect the content of these fatty acids: myristoleic acid (C14:1), palmitoleic acid (C16:1), elaidic acid (C18:1t9), linoleic acid (C18:2), and vaccenic acid (C18:1n7).

1.4. DISCUSSION

1.4.1. Growth media and initial pH

TAP and WC media previously have been used by several researchers to examine such factors as growth rate, stress response, photosynthetic rate, oil accumulation, and heavy metal removal by microalgae (Harris, 1989; Anderson, 2005; Fischer *et al.*, 2006; Jo *et al.*, 2006; Laurinavichene *et al.*, 2002; Ribeiro *et al.*, 2010; Siaut *et al.*, 2011, Vandamme *et al.*, 2011). In the present study the highest growth yield of both *Chlorella vulgaris* and *Chlamydomonas reinhardtii* was obtained using WC medium supplemented with glycyglycine. This WC medium, buffered with glycyglycine rather than Tris, uses nitrate rather than ammonium as a nitrogen source, is supplemented with vitamins and Na₂SiO₃, and contains lower concentrations of CaCl₂ and MgSO₄ than the other media tested. TAP medium originally was developed for growth of photoheterotrophs using acetate as a carbon source, and may not be optimal for photoautotrophic growth. Commonly, TAP is used for those algae which use ammonium rather than nitrate as a nitrogen source (Awasthi, 2005). BHB medium, a simple phosphate buffered salts solution formulated without a carbon source, normally is used to study hydrocarbon-degrading bacteria (Roy *et al.*, 2002) and has not previously been used for photoautotrophic growth of microalgae.

The buffering capacity of both WC medium with glycyglycine and WC medium supplemented with Tris buffer was strong as evidenced by changes in initial pH to more circumneutral values. Consistent with the known ecophysiology of these algae (Taiz, 2002), greatest growth yields were obtained under neutral to slightly alkaline conditions. Lowest growth yields occurred at the lowest (pH 4) and highest (pH 10) initial medium pH. Decreasing pH in

TAP and BHB media generally was observed with time. Fabregas (1993) noticed pronounced bacterial populations in non-axenic cultures which inhibited algal growth; however, once TRIS buffer was omitted from the medium normal microalgal growth was recovered. Ursi *et al.* (2008) found that TRIS had a deleterious effect on growth rate and pigment content of *Gracilaria birdiae*. Sofrova *et al.* (1978) demonstrated that TRIS and substances with similar structure (tricine, pentaerythritol, mannitol and sucrose) affected photochemical activity and decreased oxygen evolution and photoreduction of an artificial electron acceptor (DCPIP). Hellblom *et al.* (2001) reported a 70% reduction in the net photosynthetic rate of *Zostera marina* when TRIS buffer was used. In contrast, comparing TRIS with eight other buffers, Loeblich (1975) reported the maximal growth rate and minimal pH fluctuations when TRIS was added in a seawater medium containing vitamins and sodium nitrate as nitrogen source for growth of *Cachonina niei*. In the present study, TRIS buffer worked well for growth with WC medium but not TAP. Most likely this was due to the use of nitrate rather than ammonium as the nitrogen source in WC medium. Ammonium utilization can cause production of excess acidity. Since WC medium with glycylglycine sustained the highest yield and best buffering capacity for growth of both *Chlorella vulgaris* and *Chlamydomonas reinhardtii*, this medium was chose to conduct all subsequent experiments.

1.4.2. Growth variation at different light intensity

Photosynthetic microalgae use light as their source of energy to assimilate inorganic carbon for conversion into organic matter. In an optimal system where no other factors are limiting, light availability controls the rate of photosynthesis and cell yield (Molina-Grima *et al.*, 1999). Light requirements vary greatly among different algal species (Ojala, 1993). Photoautotrophic growth systems must be designed to optimize light availability with increasing algal density and culture depth. Growth with light intensities ranging from 125 to 1,250 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ have been reported (Lavens and Sorgeloos, 1996). In the present study, growth yield of *C. vulgaris* increased with increasing light intensity from 100 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ to 300 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$. There was no statistical difference in the observed growth yields of *Chlamydomonas reinhardtii* with light intensity. Bhola *et al.* (2011) reported that *C. vulgaris* tolerated light intensities ranging from 150 to 350 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$, but that intensities

beyond 370 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ resulted in decreased biomass production. Sorokin and Krauss (1958) reported decreased growth of *C. vulgaris* at light intensities above 90 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$. Hartig *et al.* (1998) demonstrated a linear decrease of the photochemical efficiency of PSII of unspecified microphytobenthic algae with increasing irradiances. Fischer *et al.* (2006), observed a strong negative influence on the efficiency of photosynthesis and PSII photodamage in *C. reinhardtii* at high light (2,500 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$) compared with low light (120 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$) growth conditions. They suggested that the high charge separation at PSII exceeds the availability of the electron acceptor NADP^+ at PSI resulting in the degradation of the PSII core D1 protein. Sorokin and Krauss (1958) reported decreased growth of *Chlorella pyrenoidosa* and *C. reinhardtii* at light intensities above 300 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$. Fans were used in the current study to decrease the temperature at the surface of the glass growth vessels for high light intensities. Temperatures measured at 400 to 600 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ ranged from 30 to 35°C, compared with 25 to 27°C for the experiments run at 100 to 300 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$, which may somewhat confound comparison of these data.

1.4.3. Growth variation at different CO₂ concentration

In the present study, both *C. vulgaris* and *C. reinhardtii* produced highest final yields when grown with 6% CO₂ and were capable of growth in a 12% CO₂ atmosphere. de Moraes and Costa (2007) observed no significant difference in growth rates in *Chlorella kessleri* and *Scenedesmus obliquus* cultivated with 6% and 12% CO₂. However, the growth rate value and biomass productivity of *C. kessleri* decreased when the CO₂ concentration reached 18%. Other studies have demonstrated that *C. vulgaris* can grow in elevated CO₂ atmospheres ranging from 10 to 40 % (Hirata *et al.*, 1996; Hanagata *et al.*, 1992). Maeda *et al.* (1995) found a strain of *Chlorella* sp. T-1 which could grow under 100% CO₂, even though the maximum growth rate occurred at 10%. Sung *et al.* (1999) compared the growth of a new species *Chlorella* KR-1 with concentrations of CO₂ ranging from 10 to 70% (v/v) in air. They reported optimum growth at 10% CO₂, while cultures at higher concentrations exhibited a longer lag period before growth started. Chinnasamy *et al.* (2009) reported increased chlorophyll content and biomass of *Chlorella vulgaris* ARC 1 grow with 6% and 14% CO₂.

The maximum CO₂ concentration reported for growth of *Chlamydomonas* sp. is 15% (Salih, 2011). Fischer *et al.* (2006) reported higher photosynthetic activity of *C. reinhardtii* cells

grown in a medium containing 5% CO₂ compared ambient air. Chinnasamy *et al.* (2009) reported that the carbonic anhydrase (an enzyme that catalyzes the rapid interconversion of CO₂ and water to bicarbonate and protons) of *Chlamydomonas reinhardtii* had much lower tolerance for high CO₂ levels compared with *C. vulgaris*. This mechanism may explain the inferior growth of *C. reinhardtii* at 12% compared with 3% and 6% CO₂ observed in the present study.

1.4.4. Variation of ETR at different light intensity and CO₂ concentration

The ETR measurements at different light intensities exhibited a decrease at higher light intensity (Fig. 14, 15, 16). Similar results were reported by Bhola *et al.* (2011) who noticed that for *C. vulgaris*, light intensity beyond 370 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ resulted in decreased ETR values. Also, Hartig *et al.* (1998) demonstrated a linear decrease of the actual photochemical efficiency of PSII of motile microphytobenthic algae with increasing irradiances. Fischer *et al.* (2006) showed that at high light intensity the photosynthetic electron transport chain is reduced because a high charge separation rate at PSII exceeds the availability of the electron acceptor NADP⁺ at PSI. Also, they believe that this process stimulated the degradation of the PSII core D1 protein followed by inactivation of photosynthesis and photoinhibition.

Furthermore, ETR values decreased over time. A similar observation was described by Hofstraat *et al.* (1994) who noticed higher overall ETR for *Dunaliella tertiolecta* in the first five days of the experiment compared with the observation after 17 days. Masojídek *et al.* (2001) measured the light-response curves of photosynthetic oxygen evolution and ETR of a phytoplankton mixture from a fish pond, at various times of the day. Interestingly, they observed that the highest efficiency of light utilization was at 8:30 am and the highest inhibition of photosynthesis during the day was at 10:30 am. However, at 14:00 pm the ETR increased almost to the morning value.

The overall ETR results were considerably higher for *C. reinhardtii* than *C. vulgaris* during the light or CO₂ experiments (Fig. 14, 15, 16, 18, 19, and 20). The ETR results obtained from the light experiment did not demonstrate the same rapid decrease for *C. reinhardtii* compared to *C. vulgaris*; which might indicate that *C. reinhardtii* is able to easily adapt to different light intensities. This hypothesis is supported by the presence of an “eyespot” (the simplest visual system found in nature) in flagellates like *Chlamydomonas* sp. which assist

guidance of the algae in places where light conditions are optimal for photosynthetic growth (Hegemann, 1997).

1.4.5. Total lipid content of *C. vulgaris* and *C. reinhardtii* grown at ambient, 6% and 12% CO₂

Lipid content increased with increasing CO₂ concentration. Similar findings were reported by Widjaja *et al.* (2009) who demonstrated that increasing CO₂ concentrations increased lipid productivity of *C. vulgaris*. When the CO₂ flow was increased to 50 ml/min, the lipid productivity increased by more than 50% over standard controls conditions. Fulke *et al.* (2010) found maximum lipid productivity of *Chlorella* sp. at 3% CO₂, when a range from ambient to 15% was compared. However, Chiu *et al.* (2008) found that the lipid content of *Chlorella* sp. cultured at 2%, 5%, 10%, and 15% CO₂ was very similar in their single cell photobioreactor. Production of biomass and lipid content significantly increased (6x) in a six-cell parallel photobioreactor compared with the single-cell photobioreactor, implying that detention time and light exposure were the most important factors.

The highest lipid content of total algal dry biomass was almost 18% for *C. vulgaris* and for 14% *C. reinhardtii*. These numbers are lower than have been reported by others (Demirbas 2011; Liang, 2009; Mata, 2010). In his review, Demirbas (2011) reported the average lipid content on a dry matter basis of *C. vulgaris* was between 14-22 % and for *C. reinhardtii* was 21 %. Mata (2010) reported 5-58 % lipid content for *C. vulgaris*. In our experiments, the carbon source was provided via bubbling of CO₂-enriched air into the culture medium. The reason for these low amounts of lipid content might be due to low absorption of CO₂ since the bubbles formed were large and tended to rise to surface, leading to considerable waste of gas to the atmosphere. Carvalho and Malcata (2001), studying the transfer of CO₂ into microalgal cultures, demonstrated that the use of microporous hollow fibers, rather than plain bubbling, offers technological enhancements in the effectiveness of mass transfer. In addition, this type of system offers the opportunity to recirculate the gas and to use lower gas pressures, thus reducing operating costs.

1.4.6. Fatty acids profiles of *C. vulgaris* and *C. reinhardtii* grown at ambient, 6% and 12% CO₂

The quality of biodiesel is considerably affected by the fatty acids used in its production. In the present study, GC analysis showed that the main fatty acid components of the two examined microalgae ranged from C₁₆ to C₁₈ in chain length, and generally were unsaturated. As a result, biodiesel produced from these two microalgae would have low viscosity and desirable low-temperature properties (Knothe, 2008).

The fatty acids composition of *C. vulgaris* and *C. reinhardtii* varied with different CO₂ concentrations used during growth. Palmitic acid (C16:0), stearic acid (C18:0), oleic acid (C18:1n9), linoleic acid (C18:2) were higher in content when the algae were grown at 12% CO₂. These findings are comparable to observations made by Tsuzuki *et al.* (1990). Similarly, Yoo *et al.* (2010) reported the presence palmitic acid, oleic acid, and linoleic acid as the main fatty acids in *C. vulgaris*, whereas palmitoleic acid and stearic acid were minor components. They also reported that oleic acid content was higher in *Botryococcus braunii* compared to *C. vulgaris*. Decreased content of linolenic acid (ω18:3) in *C. vulgaris* exposed to higher CO₂ concentrations also was observed by Tsuzuki *et al.* (1990). In contrast, Moreno (2011) found a significantly higher content of linolenic acid when *C. vulgaris* UTEX 259 was grown at 5% CO₂. He also observed an increase in fatty acids grown at higher CO₂ concentration compared with ambient.

1.5. CONCLUSIONS

The choice of microalgae for CO₂ mitigation and production of biodiesel requires selection according to their growth rate, lipid content, and tolerance of high levels of CO₂. Results showed that *C. vulgaris* had a faster growth rate, higher lipid content, and better tolerance for high CO₂ levels compared to *C. reinhardtii* (Table 6, Table 11, Fig. 15). Furthermore, data indicate that CO₂ concentrations result in higher lipid content.

Wright's Cryptophytes medium supplemented with glycylglycine (WC-Gly) had the best buffering capacity and the highest growth rate among the media tested: Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes medium supplemented with TRIS (WC-Tris). A light intensity of 200 μmol photons m⁻²s⁻¹ provided the best growth yield and ETR.

In future studies, it would be valuable to examine the effect of different temperature on the growth yield and ETR. After nutrient quantity and quality, environmental temperature is the key parameter that controls the basic rates of all biochemical reactions in the algal cell. By optimizing the temperature, it may be possible to increase the biomass productivity and lipid content can be increased as has been suggested by Sandnes *et al.*, 2005, Converti *et al.*, 2009.

Examination on possible nitrogen limitation effects on the lipid content and biomass of the microalgae could provide a better perspective on the best method for enhancing algae biofuel. Deng *et al.* (2011) noticed that the external carbon source has little impact on lipid accumulation whereas N or S deficiency in a high carbon medium achieved significantly higher lipid accumulation.

1.6. REFERENCES

- Aksoy F, Bayrakceken H, Eryilmaz T, Aksoy L.** 2011. Analyzing the impact of using different methyl esters in a diesel engine on engine performance and emissions. *Energy Educ Sci Technol Part A.* **27:** 25–34.
- Allen EJ, Nelson, EW.** 1910. On the artificial culture of marine plankton organisms. *J. Mar. Biol. Assoc. U.K.* **8:** 421-474.
- Anderson A.** 2005. *Algal culturing technique.* Elsevier Academic Press, Burlington. ISBN 0-12-088426-7.
- Awasthi A.** 2005. Nitrate reductase activity: A solution to nitrate problems tested in free and Immobilized algal cells in presence of heavy metals. *Int. J. Environ. Sci. Tech.* **2:** 201-206.
- Axelsson L, Mercado JM, Figueroa FL.** 2000. Utilization of HCO_3^- at high pH by the brown macroalga *Laminaria saccharina*. *Eur. J. of Phycol.* **35:** 53-59.
- Balat H.** 2010. Prospects of biofuels for a sustainable energy future: a critical assessment. *Energy Educ Sci Technology Part A.* **24:**85-111.
- Beneman J R, Tillet DM, Weissman JC.** 1987. Micro-algae Biotechnology. *Trends in Biotechnology.* **5:** 47-53.
- Bhola V, Desikan R, Santosh SK, Subburamu K, Sanniyasi E, Bux F.** 2011. Effects of parameters affecting biomass yield and thermal behaviour of *Chlorella vulgaris*. *J. Biosci. Bioeng.* **111:** 377-382.
- Bouterfas R, Belkoura M, Dauta A.** 2006. The effects of irradiance and photoperiod on the growth rate of three freshwater green algae isolated from a eutrophic lake. *Limnetica,* **25:** 647-656.
- Burlew J.** 1976. *Algae culture from laboratory to pilot plant.* Carnegie Institution of Washington Publication 600. Washington. D.C.
- Carvalho AP, Malcata FX.** 2001. Transfer of carbon dioxide within cultures of microalgae: plain bubbling versus hollow-fiber modules. *Biotechnol Prog.* **17:**265-72.
- Chini ZG, Tomasello V, Pinzani E, Tredici MR.** 1996. Outdoor cultivation of *Arthrospira platensis* during autumn and winter in temperate climates. *J Appl Phycol.* **8:**293–301.
- Chinnasamy S, Ramakrishnan B, Bhatnagar A, Das KC.** 2009. Biomass production potential of a wastewater alga *Chlorella vulgaris* ARC 1 under elevated levels of CO_2 and temperature. *Int. J. Mol. Sci.* **10:** 518-532.

Chiu S-Y, Kao C-Y, Chen C-H, Kuan T-C, Ong S-C, Lin C-S. 2008. Reduction of CO₂ by a high-density culture of *Chlorella* sp. in a semicontinuous photobioreactor. *Bioresource Technology*. **99**: 3389-3396.

Converti A, Casazza AA, Ortiz EY, Perego P, Del Borghi M. 2009. Effect of temperature and nitrogen concentration on the growth and lipid content of *Nannochloropsis oculata* and *Chlorella vulgaris* for biodiesel production. *Chemical Engineering and Processing*. **48**: 1146–1151.

Croft MT, Warren MJ, Smith AG. 2006. Algae Need Their Vitamins. *Eukaryotic Cell*. **5**: 1175–1183.

Demirbas MF. 2009. Biorefineries for biofuel upgrading: a critical review. *Appl Energy*. **86**: S151–S161.

Demirbas MF. 2011. Biofuels from algae for sustainable development. *Applied Energy*. **88**: 3473-3480.

Eriksen NT. 2008. The technology of microalgal culturing. *Biotechnol Lett*. **30**:1525–153.

Deng X, Fei X, Li Y. 2011. The effects of nutritional restriction on neutral lipid accumulation in *Chlamydomonas* and *Chlorella*. *African Journal of Microbiology Research*. **5**: 260-270.

Fabregas J, Vazquez V, Cabezas B, Otero A. 1993. TRIS not only controls the pH in microalgal cultures, but also feeds bacteria. *J. Appl. Phycol*. **5**:543–5.

Fischer BB, Wiesendanger M, Rik I.L.E. 2006. Growth condition-dependent sensitivity, photodamage and stress response of *Chlamydomonas reinhardtii* exposed to high light conditions. *Plant Cell Physiol*. **47**: 1135–1145.

Foy RH, Gibson CE. 1993. The influence of irradiance, photoperiod and temperature on the growth kinetic of three planktonic diatoms. *Eur. J. Phycol*. **28**: 203-212.

Francisco EC, Neves DB, Jacob-Lopes E, Franco TT. 2010. Microalgae as feedstock for biodiesel production: Carbon dioxide sequestration, lipid production and biofuel quality. *J Chem Technol Biotechnol*. **85**: 395-403.

Fulke AB, Mudliar SN, Yadav R, Shekh A, Srinivasan N, Ramanan R, Krishnamurthi K, Devi SS, Chakrabarti T. 2010. Bio-mitigation of CO₂, calcite formation and simultaneous biodiesel precursors production using *Chlorella* sp. *Bioresour Technol*. **101**: 8473-8476.

Gorman DS, Levine RP. 1965. Cytochrome f and plastocyanin: their sequence in the photosynthetic electron transport chain of *Chlamydomonas reinhardtii*. *Proc. Natl. Acad. Sci*. **54**: 1665-1669.

Hanagata N, Takeuchi T, Fukuju Y. 1992. Tolerance of Microalgae to High CO₂ and High Temperature. *Phyto-chemistry*. **10**: 3345-3348.

- Hartig P, Wolfstein K, Lippenmeier S, Colijn F.** 1998. Photosynthetic activity of natural microphytobenthos populations measured by fluorescence (PAM) and ^{14}C -tracer methods: a comparison. *Mar. Ecol. Prog. Ser.* **166**: 53–62.
- Harris EH.** 1989. *The Chlamydomonas sourcebook. A comprehensive guide to biology and laboratory use.* San Diego: Academic Press Inc.
- Hegemann P.** 1997. Vision in microalgae. *Planta.* **203**: 265-274.
- Hellblom F, Beer S, Björk M, Axelsson L.** 2001. A buffer sensitive inorganic carbon utilization system in *Zostera marina*. *Aquatic Botany.* **69**: 55-62.
- Hirata S, Hayashitani M, Taya M, Tone S.** 1996. Carbon dioxide fixation in batch culture of *Chlorella* sp. using a photobioreactor with a sunlight-collection device. *J. Ferment. Biobng.* **5**: 470-472.
- Hobson LA, Hartley FA, DE Ketcham.** 1979. Effects of variations in daylength and temperature on net rates of photosynthesis, dark respiration and excretion by *Isochrysis galbana* Parke. *Plant Physiol.* **63**: 947-951.
- Hodson RC, Williams II SK, Davidson WR Jr.** 1975. Metabolic control of urea catabolism in *Chlamydomonas reinhardi* and *Chlorella pyrenoidosa*. *J of Bacteriology.* **3**:1022-1035.
- Hofstraat JW, Peeters JCH, Snel JFH, Geel C.** 1994. Simple determination of photosynthetic efficiency and photoinhibition of *Dunaliella tertiolecta* by saturating pulse measurements. *Mar. Ecol. Prog. Ser.* **103**: 187–196.
- Jo JH, Lee DS, Park JM.** 2006. Modeling and optimization of photosynthetic hydrogen gas production by green alga *Chlamydomonas reinhardtii* in sulfur-deprived circumstance. *Biotechnol. Prog.* **22**: 431-437.
- Knothe G.** 2008. “Designer” biodiesel: optimizing fatty ester composition to improve fuel properties. *Energy Fuels.* **22**: 1358–1364.
- Laurinavichene TV, Tolstygina IV, Galiulina RR, Ghirardi ML, Seibert M, Tsygankov AA.** 2002. Dilution methods to deprive *Chlamydomonas reinhardtii* cultures of sulfur for subsequent hydrogen photoproduction. *Intern J of Hydrogen Energ.* **27**: 1245-1249.
- Laval D, Mazliak PM.** 1995. *Nutrition et métabolisme (Physiologie végétale).* Ed. Hermann. Paris. 539 pp.
- Lavens P, Sorgeloos P.** 1996. *Manual on the production and use of live food for aquaculture.* Fisheries and Aquaculture Department. FAO Corporate Document Repository.
<http://www.fao.org/docrep/003/W3732E/w3732e00.HTM>

- Liang Y, Sarkany N, Cui Y.** 2009. Biomass and lipid productivities of *Chlorella vulgaris* under autotrophic, heterotrophic and mixotrophic growth conditions. *Biotechnol Lett.* **31**:1043-9.
- Loeblich A.** 1975. A seawater medium for dinoflagellates and the nutrition of *Cachonina niei*. *J. Phycol.* **11**:80– 86.
- Ma N, Chalmers JJ, Aunins JG, Zhou W, Xie L.** 2004. Quantitative studies of cell–bubble interactions and cell damage at different Pluronic F-68 and cell concentrations. *Biotechnol Prog.* **20**:1183–1191.
- Maeda K, Owada M, Kimura N, Omata L, Karube I.** 1995. CO₂ fixation from the flue gas on coalfired thermal power plant by microalgae. *Energy Convers. Manage.* **36**: 717-720.
- Mata TM, Antonio AM, Nidia SC.** 2010. Microalgae for biodiesel production and other applications: A review. *Renewable and Sustainable Energy Reviews.* **14**:217–232.
- Masojídek J, Grobbelaar JU, Pechar L, Koblížek M.** 2001. Photosystem II electron transport rates and oxygen production in natural waterblooms of freshwater cyanobacteria during a diel cycle. *J. Plankton Res.* **23**:57-66.
- McLachlan J.** 1973. Growth media-marine. In: Stein, J., ed. *Handbook of Phycological Methods: Culture Methods and Growth Measurements.* Cambridge University Press, Cambridge, UK, pp. 25–51.
- Milner HW.** 1949. The fatty acids of *Chlorella*. *Jour. Biol. Chem.* **176**: 813-817.
- Molina-Grima EM, Robles Medina A, Gimenez Gimenez A, Sanchez Perez JA, Garcia Camacho F, Garcia Sanchez JL.** 1994. Comparison between extraction of lipids and fatty acids from microalgal biomass. *J Am Oil Chem Soc.* **71**:955 –959.
- Molina-Grima EM, Fernandez FGA, Camacho FG, Christi Y.** 1999. Photobioreactors: light regime, mass transfer, and scaleup. *J. Biotechnol.* **70**:231–247.
- de Morais MG, Costa JAV.** 2007. Isolation and selection of microalgae from coal fired thermoelectric power plant for biofixation of carbon dioxide. *Energy Convers. Manage.* **48**: 2169–2173.
- Moreno R.** 2011. Identification of algal strains by PCR amplification and evaluation of their fatty acid profiles for biodiesel production. Master’s Thesis. Louisiana State University.
- Ojala A.** 1993. Effects of temperature and irradiance on growth of two freshwater photosynthetic cryptophytes. *J. Phycol.* **29**: 278-284.
- Piquemal F.** 1990. Contribution à l’étude ecophysiologique de quelques especes d’algues en culture. These de Doctorat, Universite Paul Sabatier, Toulouse, 137 pp.

- Provasoli L, Carlucci AF.** 1974. Vitamins and growth regulators. In: Stewart, W. D. P., ed. *Algal Physiology and Biochemistry*. Blackwell Scientific, UK, pp. 741–87.
- Rados S, Vaclav B, Frantisek D.** 1975. CO₂ Balance in industrial cultivation of algae. *Arch Hydrobiol.* **46**:297–310.
- Redalje DG, Laws EA.** 1983. The effect of environmental factors on growth and the chemical and biochemical composition of marine diatoms 1-light and temperature effects. *J. Exp. Mar. Biol. Ecol.* **68**: 59-79.
- Ribeiro RFL, Magalhães SMS, Barbosa FAR, Nascentes CC, Campos IC, Moraes DC** 2010. Evaluation of the potential of microalgae *Microcystis novacekii* in the removal of Pb²⁺ from an aqueous medium. *J of Hazard Mat.* **179**:947–953.
- Richmond A.** 2004. *Handbook of Microalgal culture: Biotechnology and Applied phycology*. Blackwell publishing, Oxford, UK.
- Rickert KW, Sears J, Beck WF, Brudvig GW.** 1991. Mechanism of irreversible inhibition of O₂ evolution in photosystem-II by TRIS (hydroxymethyl)-aminomethane. *Biochem.* **30**: 7888-7894.
- Riebesell U, Wolfgladrow DA, Smetacek V.** 1993. Carbon dioxide limitation of marine phytoplankton growth rates. *Nature.* **361**:249–51.
- Rodolfi L, Chini Zittelli G, Bassi N, Padovani G, Biondi N, Bonini G, Tredic MR.** 2009. Microalgae for oil: strain selection, induction of lipid synthesis and outdoor mass cultivation in a low-cost photobioreactor. *Biotech and Bioengineering.* **102**: 100–112.
- Roy S, Hens D, Biswas D, Biswas D, Kumar R.** 2002. Survey of petroleum - degrading bacteria in coastal waters of Sunderban. Biosphere Reserve. *World J Microbiol Biotech.* **18**:575-581.
- Salih FM.** 2011. Microalgae Tolerance to High Concentrations of Carbon Dioxide: A Review. *Journal of Environmental Protection.* **2**: 648-654.
- Sandnes JM, Källqvist T, Wenner D, Gislerød HR.** 2005. Combined influence of light and temperature on growth rates of *Nannochloropsis oceanica*: linking cellular responses to large-scale biomass production. *Journal of Applied Phycology.* **17**: 515–525.
- Schenk PM, Thomas-Hall SR, Stephens E, Marx UC, Mussgnug JH, Posten C, Kruse O, Hankamer B.** 2008. Second generation biofuels: high-efficiency microalgae for biodiesel production. *Bioenergy. Res.* **1**: 20–43.
- Siaut M, Cuiné S, Cagnon C, Fessler B, Nguyen M, Carrier P, Beyly A, Beisson F, Triantaphylidès C, Li-Beisson Y, Peltier G.** 2011. Oil accumulation in the model green alga

Chlamydomonas reinhardtii: characterization, variability between common laboratory strains and relationship with starch reserves. BMC Biotechnology. **11**: 7-21.

Sofrova D, Wilhelm J, Naus J, Leblová S. 1978. Effects of TRIS and analogous hydroxycompounds on photosystem II of blue-green algae. Photosynthetica. **12**: 391-398.

Spoehr HA, Milner HW. 1949. The chemical composition of *Chlorella*; effect of environmental conditions. Plant Physiol. **24**: 120-149.

Sorokin C, Krauss RW. 1965. The dependence of cell division in *Chlorella* on temperature and light intensity. American J of Botany. **52**: 331-339.

Sung K-D, Lee J-S, Shin C-S, Park S-C, Choi M-J. 1999. CO₂ fixation by *Chlorella* sp. KR-1 and its cultural characteristics. Bioresource Technology. **68**: 269-273.

Taiz L, Zeiger E. 2002. Plant Physiology, 3rd ed. Sinauer Associates.

Tsuzuki M, Ohnuma E, Norihiro S, Tadashi T, Akihiko K. 1990. Effects of CO₂ concentration during growth on fatty acid composition in microalgae. Plant Physiol. **93**: 851-856.

Ursi S, Guimarães M, Plastino EM. 2008. Deleterious effect of TRIS buffer on growth rates and pigment content of *Gracilaria birdiae* Plastino & E.C. Oliveira (Gracilariales, Rhodophyta). Acta bot. bras. **22**: 891-896.

Vandamme D, Vieira Pontes SC, Goiris K, Foubert I, Jan Pinoy LJ, Muylaert K. 2011. Evaluation of electro-coagulation–flocculation for harvesting marine and freshwater microalgae. Biotechnology and Bioengineering. **108**: 2320–2329.

Widjaja A, Chien C-C, Yi-Hsu J. 2009. Study of increasing lipid production from fresh water microalgae *Chlorella vulgaris*. Journal of the Taiwan Institute of Chemical Engineers. **40**: 13–20.

Yoo C, Jun S-Y, Lee J-Y, Ahn C-Y, Oh H-M. 2010. Selection of microalgae for lipid production under high levels carbon dioxide. Bioresource Technology. **101**: S71–S7.

Yun Y, Park JM. 1997. Development of gas recycling photobioreactor system for microalgal carbon dioxide fixation. Korean J Chem Eng. **14**: 297–300.

Table 1. Chemical composition of TAP medium (Gorman and Levine, 1965).

Stock Solution (SL)	Quantity	Component	Concentration in SL	Conc. in final Medium
Tris base	2.42 g	H ₂ NC(CH ₂ OH) ₃ Tris(hydroxymethyl)-aminomethan		2.00 x 10 ⁻² M
TAP-salts (Beijerinck salts)	25 mL	NH ₄ Cl	15 g L ⁻¹	7.00 x 10 ⁻³ M
		MgSO ₄ • 7H ₂ O	4 g L ⁻¹	8.30 x 10 ⁻⁴ M
		CaCl ₂ • 2H ₂ O	2 g L ⁻¹	4.50 x 10 ⁻⁴ M
Phosphate solution	0.375 mL	K ₂ HPO ₄	28.8 g 100 mL ⁻¹	1.65 x 10 ⁻³ M
		KH ₂ PO ₄	14.4 g 100 mL ⁻¹	1.05 x 10 ⁻³ M
Trace elements solution (Hutner trace elements)	1 mL	Na ₂ EDTA • 2H ₂ O	5.00 g 100 mL ⁻¹	1.34 x 10 ⁻⁴ M
		ZnSO ₄ • 7H ₂ O	2.20 g 100 mL ⁻¹	1.36 x 10 ⁻⁴ M
		H ₃ BO ₃	1.14 g 100 mL ⁻¹	1.84 x 10 ⁻⁴ M
		MnCl ₂ • 4H ₂ O	0.50 g 100 mL ⁻¹	4.00 x 10 ⁻⁵ M
		FeSO ₄ • 7H ₂ O	0.50 g 100 mL ⁻¹	3.29 x 10 ⁻⁵ M
		CoCl ₂ • 6H ₂ O	0.16 g 100 mL ⁻¹	1.23 x 10 ⁻⁵ M
		CuSO ₄ • 5H ₂ O	0.16 g 100 mL ⁻¹	1.00 x 10 ⁻⁵ M
		(NH ₄) ₆ MoO ₂₄ • 4H ₂ O	0.11 g 100 mL ⁻¹	4.44 x 10 ⁻⁶ M

Table 2. Chemical composition of WC medium (Guillard and Lorenzen, 1972).

Component	Quantity	Stock Solution	Molar Concentration in Final Medium
Buffer: Glycylglycine Tris base	500 mg	---	3.78×10^{-3} M
NaNO ₃	1 mL	85.01 g L ⁻¹ dH ₂ O	1.00×10^{-3} M
CaCl ₂ • 2H ₂ O	1 mL	36.76 g L ⁻¹ dH ₂ O	2.50×10^{-4} M
MgSO ₄ • 7H ₂ O	1 mL	36.97 g L ⁻¹ dH ₂ O	1.50×10^{-4} M
NaHCO ₃	1 mL	12.60 g L ⁻¹ dH ₂ O	1.50×10^{-4} M
Na ₂ SiO ₃ • 9H ₂ O	1 mL	28.42 g L ⁻¹ dH ₂ O	1.00×10^{-4} M
K ₂ HPO ₄	1 mL	8.71 g L ⁻¹ dH ₂ O	5.00×10^{-5} M
Trace metal solution	1 mL	(see below)	---
Vitamin solution	1 mL	(see below)	---

Table 3. Trace metal solution for WC medium.

Component	Primary Stock Solution	Quantity	Molar Concentration in Final Medium
FeCl ₃ • 6H ₂ O	---	3.15 g	1.17 x 10 ⁻⁵ M
Na ₂ EDTA • 2H ₂ O	---	4.36 g	1.17 x 10 ⁻⁵ M
CuSO ₄ • 5H ₂ O	10.0 g L ⁻¹ dH ₂ O	1 mL	4.01 x 10 ⁻⁸ M
Na ₂ MoO ₄ • 2H ₂ O	6.0 g L ⁻¹ dH ₂ O	1 mL	2.48 x 10 ⁻⁸ M
ZnSO ₄ • 7H ₂ O	22.0 g L ⁻¹ dH ₂ O	1 mL	7.65 x 10 ⁻⁸ M
CoCl ₂ • 6H ₂ O	10.0 g L ⁻¹ dH ₂ O	1 mL	4.20 x 10 ⁻⁸ M
MnCl ₂ • 4H ₂ O	180.0 g L ⁻¹ dH ₂ O	1 mL	9.10 x 10 ⁻⁷ M
H ₃ BO ₃	---	1.00 g	1.62 x 10 ⁻⁵ M

Table 4. Vitamin solution for WC medium.

Component	Primary Stock Solution	Quantity	Molar Concentration in Final Medium
Thiamine • HCl (vit. B ₁)	---	100 mg	2.96 x 10 ⁻⁷ M
Biotin (vit. H)	0.5 g/L dH ₂ O	1 mL	2.05 x 10 ⁻⁹ M
Cyanocobalamin (vit. B ₁₂)	0.5 g/L dH ₂ O	1 mL	3.69 x 10 ⁻¹⁰ M

Figure 1. The experimental set-up for evaluation of growth medium and pH (before inoculation). The mouths of the tubes were covered with foam stoppers. Air at 50 mL/min was bubbled through the testing tubes using a pipette to agitate the solution. A syringe provides an exit/sampling point.



Figure 2. The experimental set-up for evaluation of growth medium and pH (day 5). Visible growth at pH 7 and pH 8. Controls show no growth.



Figure 3. The experimental set-up for optimization of light intensity. Test tubes were placed on their sides on two shelves illuminated with 460-watt metal GE grow lights source. Light was varied by rheostat from 100-600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ (12 hours light/ 12 hours dark).

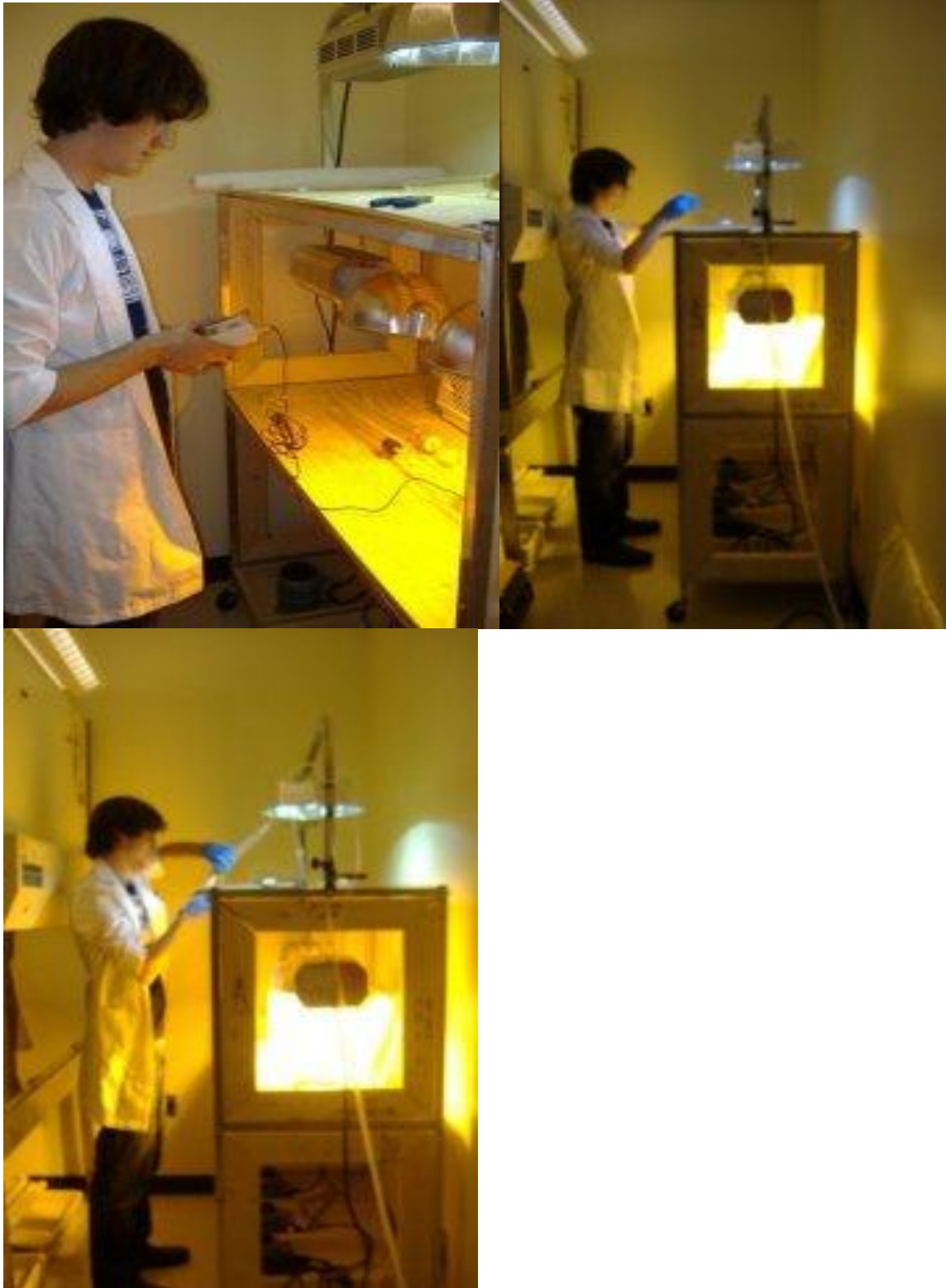


Figure 4. Phytoplankton Analyzer PHYTO-PAM (Heinz Walz GmbH, Germany) was used to measure the relative electron transport rate (ETR) as a function of PAR (Photosynthetically Active Radiation) on day two, three and four.



Figure 5. The experimental set-up for evaluation of CO₂ concentration. Glass serum bottles filled with WC medium and sealed with solid stoppers. Two needles were inserted in the stopper to bubble the gas (CO₂ level=0.038%, 3%, 6%, and 12%) through the medium and to provide an exit/sampling point. Light was held constant at 200 $\mu\text{mol m}^{-2}\text{s}^{-1}$. The medium was stirred continuously using magnetic stirrers.



Fig 6. 8000M Mixer/Mill® system (Spex SamplePrep,USA).



Fig.7. Soxtec™ 2055 system (Foss Analytical, Denmark).



Figure 8. *Chlorella vulgaris* growth curve using Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC) medium over a 5-day period. The initial pH of the media ranged from pH 4 to pH 10.

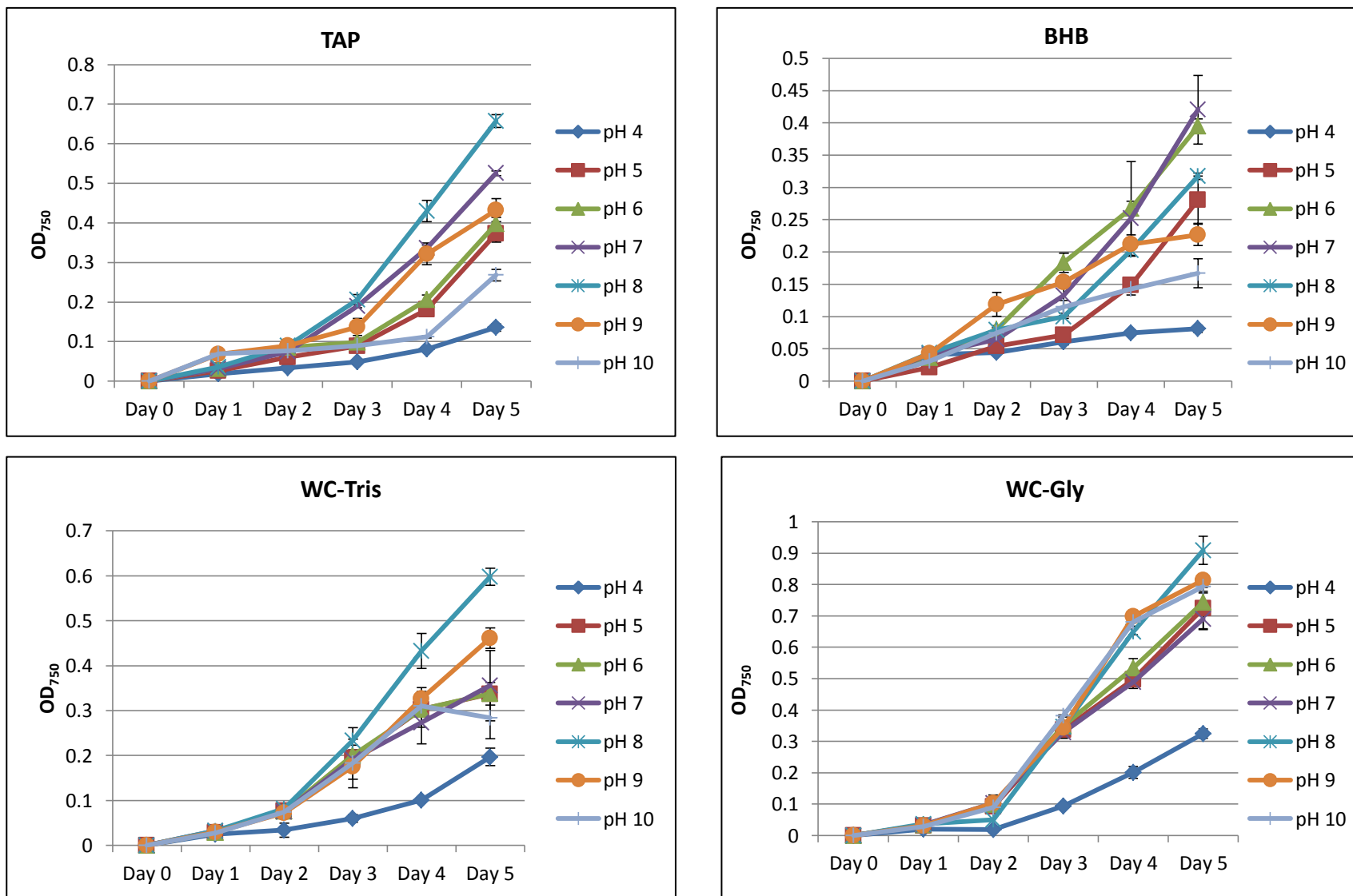


Figure 9. pH variation during growth of *Chlorella vulgaris* growth using Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC) medium over a 5-day period.

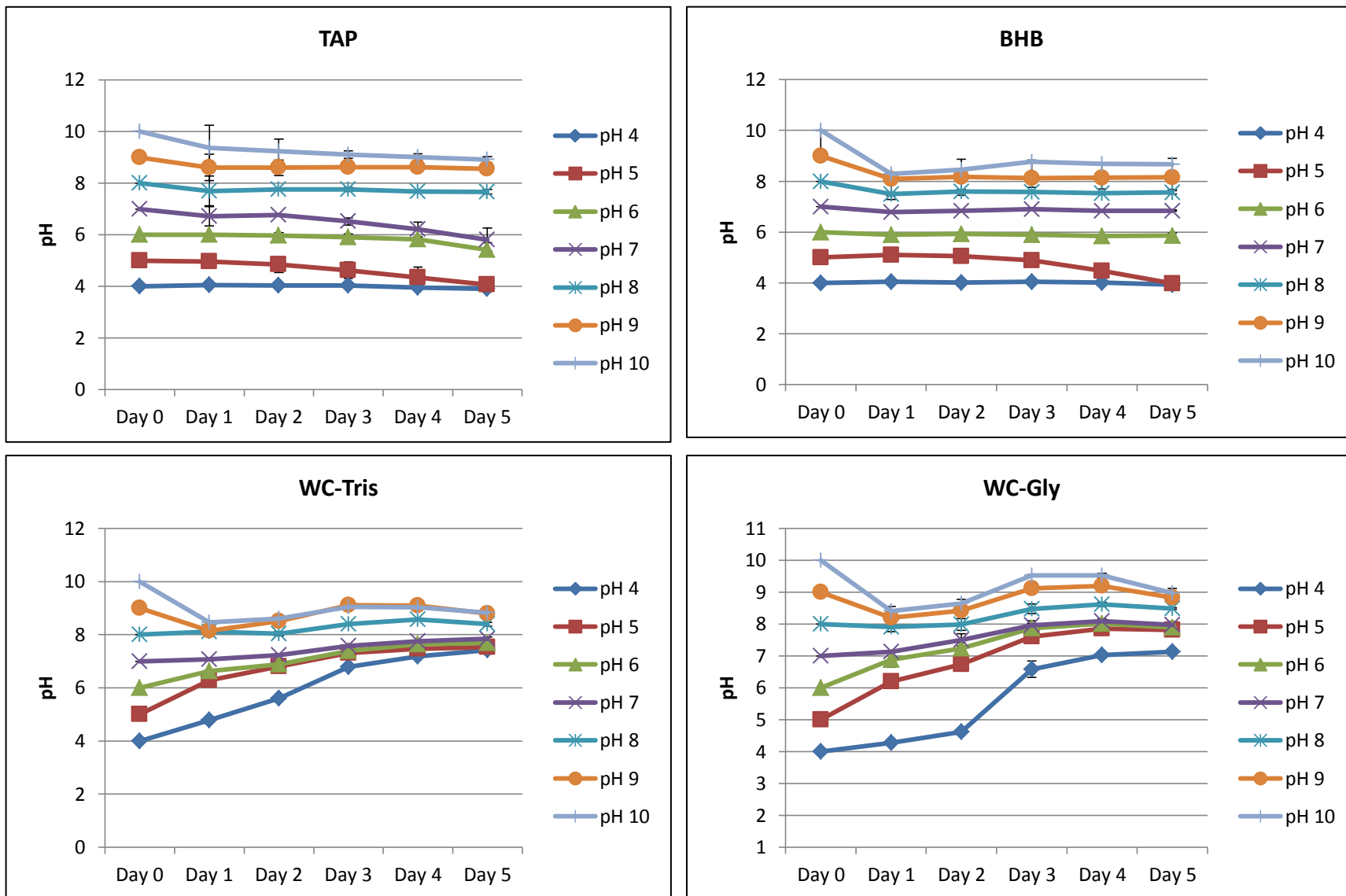


Figure 10. Growth of *Chlamydomonas reinhardtii* growth curve using Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC) medium over a 5-day period. The initial pH of the media ranged from pH 4 to pH 10.

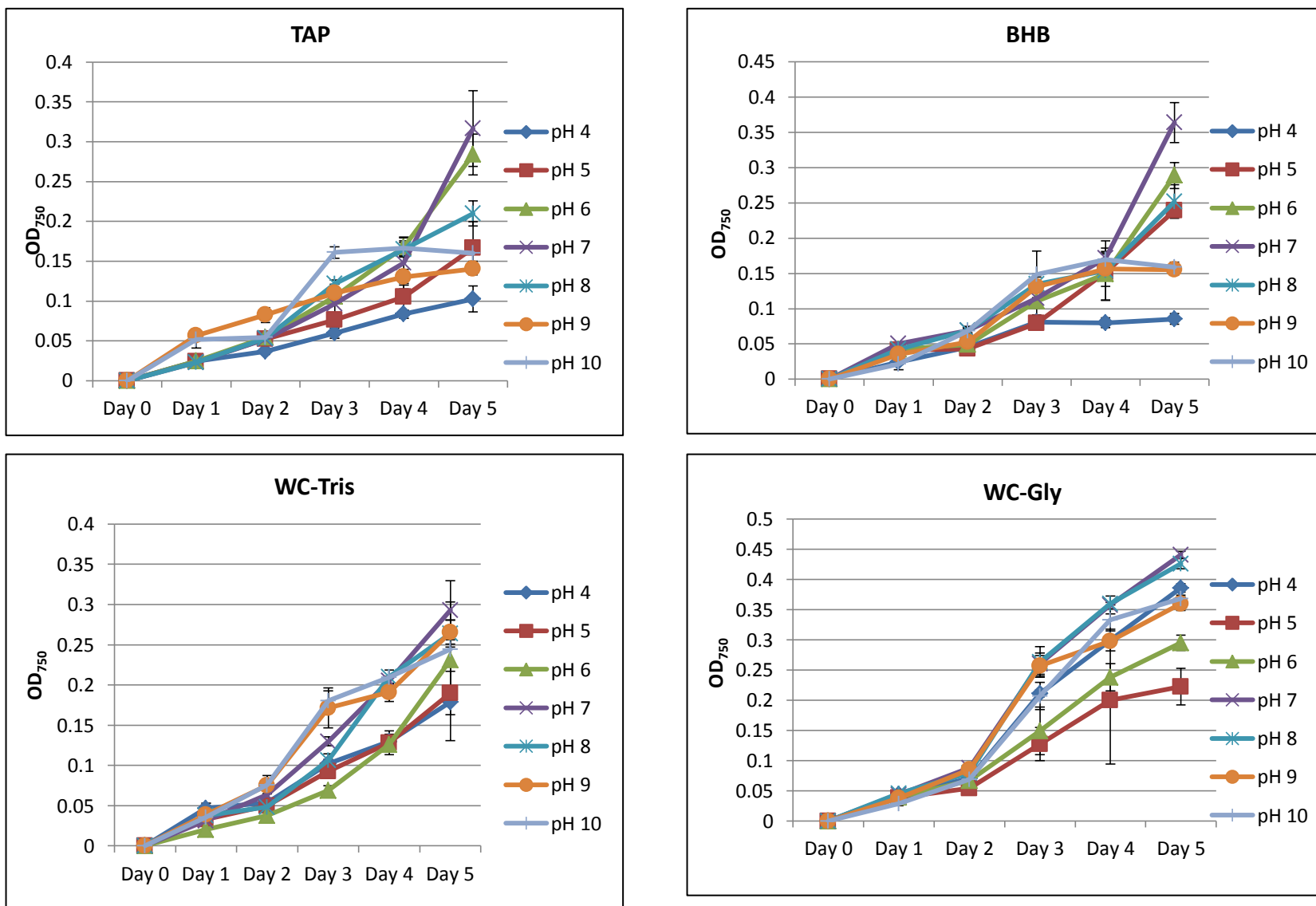


Figure 11. pH variation during *Chlamydomonas reihardtii* growth using Tris Acetate Phosphate (TAP), Bushnell Haas Broth (BHB), and Wright's Cryptophytes (WC) medium over a 5-day period.

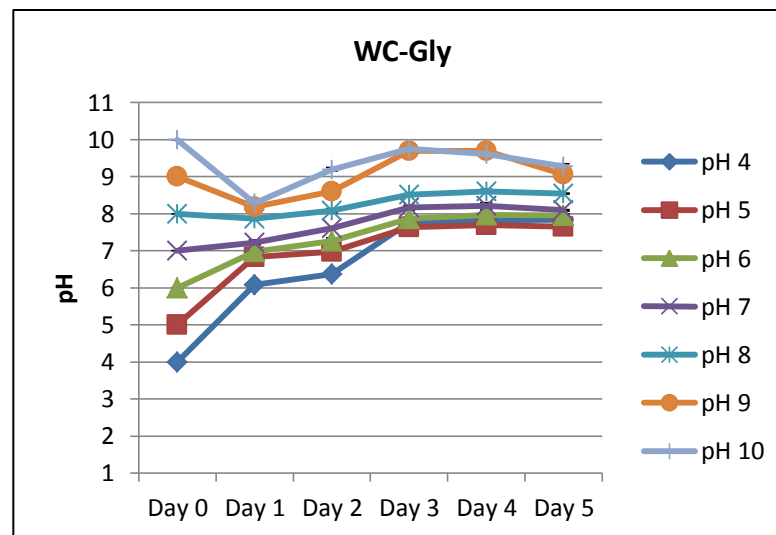
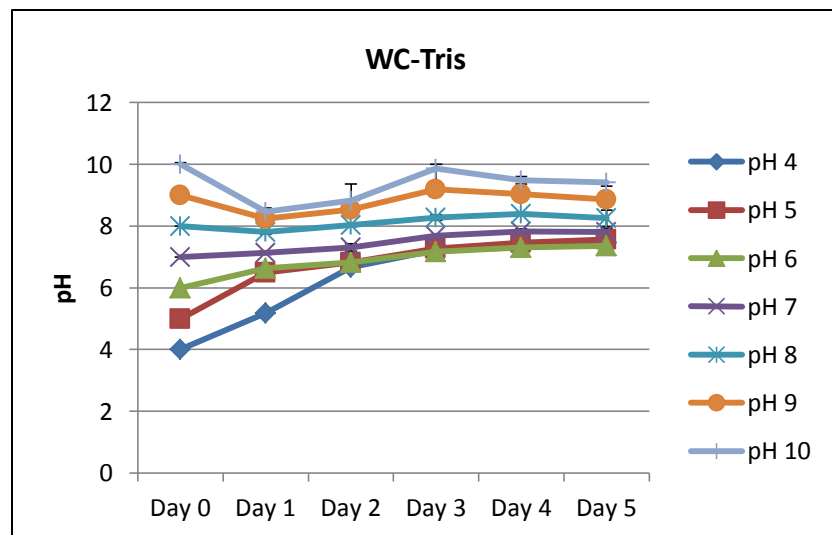
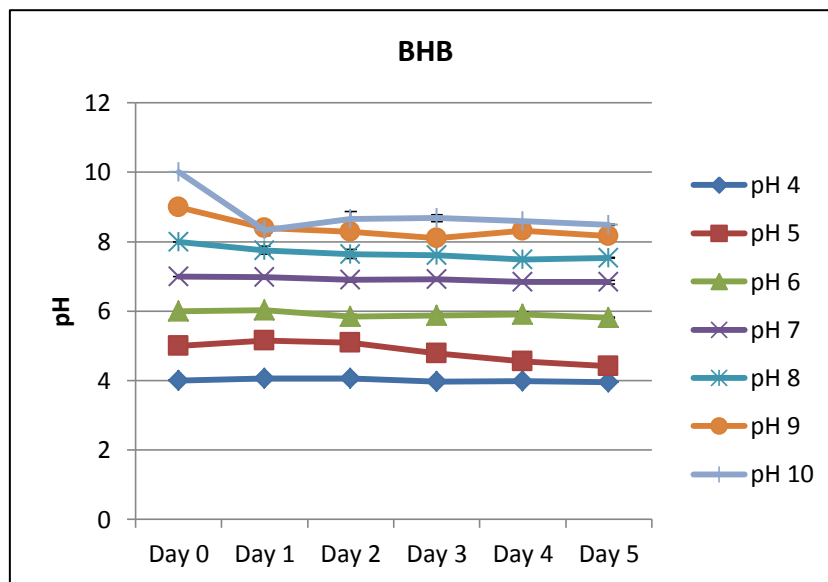
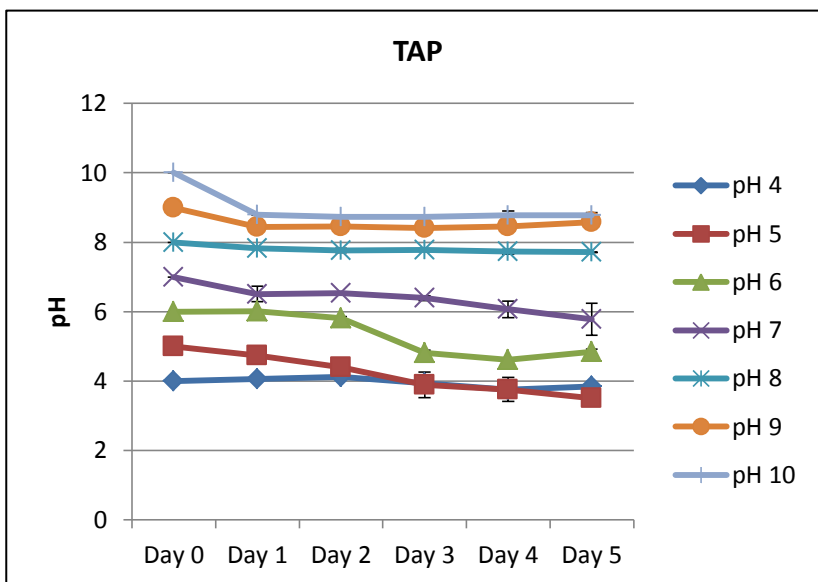


Fig. 12. Variation of pH in un-inoculated vessels.

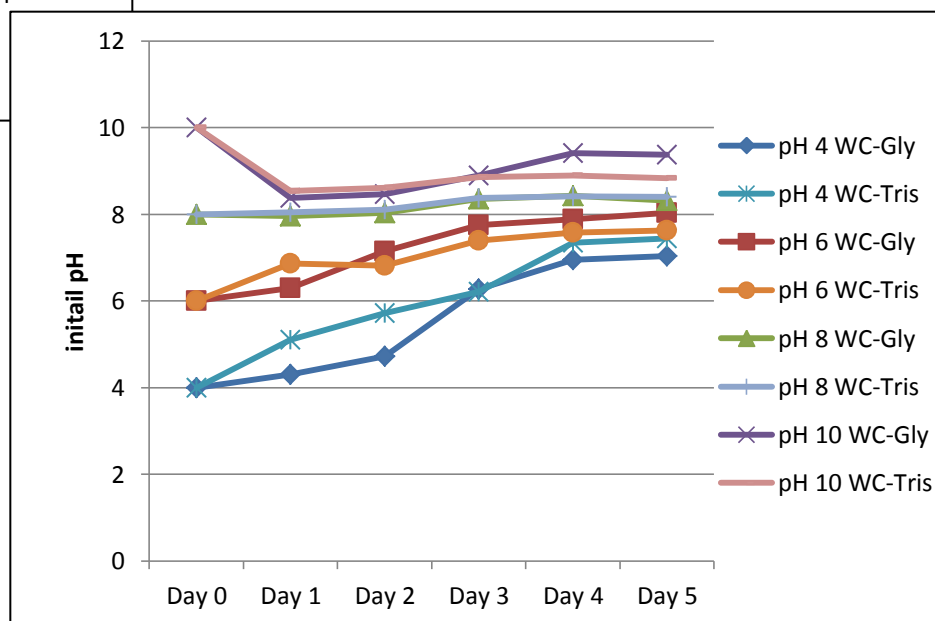
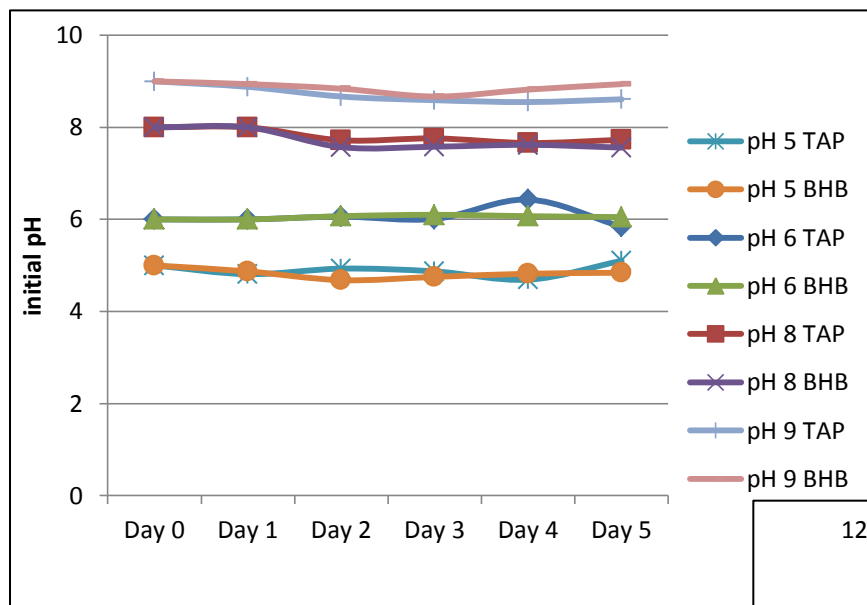


Table 5. Analysis of variance (ANOVA) table at day five for TAP and BHB medium.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	27	2.0287	150.2228	<.0001
Algae type	1	0.3962	792.2854	<.0001
Medium type	1	0.0626	125.1723	<.0001
Algae type* Medium type	1	0.1725	344.9751	<.0001
Initial pH	6	1.1010	366.8946	<.0001
Algae type* Initial pH	6	0.1258	41.9232	<.0001
Medium type * Initial pH	6	0.0498	16.5848	<.0001
Algae type* Medium type* Initial pH	6	0.1206	40.1945	<.0001
Error	84	0.0420		
Total	111	2.0708		

Table 6. Comparison of OD₇₅₀ at day five between the algae species (*=significant) and among all pH tested (Levels not connected by same letter are significantly different).

Medium type	TAP		BHB		WC-Gly		WC-Tris	
	<i>C. vulgaris</i>	<i>C. reinhardtii</i>	<i>C. vulgaris</i>	<i>C. reinhardtii</i>	<i>C. vulgaris</i>	<i>C. reinhardtii</i>	<i>C. vulgaris</i>	<i>C. reinhardtii</i>
Initial pH								
4	0.135 ^f	0.102 ^d	0.080 ^e	0.087 ^d	0.324 ^e	0.386 ^{*b}	0.197 ^d	0.223 ^{bc}
5	0.371 ^{*d}	0.170 ^c	0.269 ^c	0.237 ^{bc}	0.707 ^{*d}	0.229 ^d	0.332 ^{*bc}	0.190 ^c
6	0.390 ^{*d}	0.283 ^a	0.395 ^{*a}	0.289 ^{ab}	0.746 ^{*cd}	0.298 ^c	0.341 ^{*bc}	0.233 ^{bc}
7	0.512 ^{*b}	0.318 ^a	0.423 ^a	0.367 ^a	0.688 ^{*d}	0.439 ^a	0.382 ^{*b}	0.286 ^a
8	0.652 ^{*a}	0.207 ^b	0.314 ^b	0.299 ^{ab}	0.910 ^{*a}	0.442 ^a	0.582 ^{*a}	0.265 ^{ab}
9	0.431 ^{*c}	0.137 ^{cd}	0.227 ^{*c}	0.155 ^d	0.832 ^{*b}	0.361 ^b	0.505 ^{*a}	0.263 ^{ab}
10	0.268 ^{*e}	0.160 ^c	0.169 ^d	0.161 ^{cd}	0.796 ^{*bc}	0.365 ^b	0.282 ^{cd}	0.258 ^{ab}

Table 7. Comparison between initial pH and final pH (*=significant different) ($\alpha = 0.05$).

A) *C. vulgaris*

Medium type	Final pH			
Initial pH	TAP	BHB	WC-Gly	WC-Tris
4	3.90	3.90	7.14*	7.41*
5	4.00*	3.97*	7.81*	7.52*
6	5.47*	5.79*	7.89*	7.67*
7	5.90*	6.82*	7.95*	7.84*
8	7.67*	7.53*	8.47*	8.38*
9	8.56*	8.14*	8.84*	8.76*
10	8.92*	8.66*	8.95*	8.76*

B) *C. reinhardtii*

Medium type	Final pH			
Initial pH	TAP	BHB	WC-Gly	WC-Gly
4	3.85*	3.93	7.83*	7.48*
5	3.53*	4.44*	7.67*	7.55*
6	4.84*	5.82*	7.91*	7.40*
7	5.79*	6.81*	8.09*	7.81*
8	7.74*	7.54*	8.53*	8.29*
9	8.57*	8.13*	8.99	8.87*
10	8.78*	8.47*	9.25*	9.43*

Table 8. Analysis of variance (ANOVA) table at day five for WC-Gly and WC-Tris medium.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	27	4.8698	248.7527	<.0001
Algae type	1	1.6391	2260.5490	<.0001
Medium type	1	1.4476	1996.4610	<.0001
Algae type* Medium type	1	0.3563	491.3842	<.0001
Initial pH	6	0.7150	164.3561	<.0001
Algae type* Initial pH	6	0.5144	118.2521	<.0001
Medium type * Initial pH	6	0.0605	13.9157	<.0001
Algae type* Medium type* Initial pH	6	0.1369	31.4642	<.0001
Error	84	0.0609		
Total	111	4.9307		

Table 9. Analysis of variance (ANOVA) table of OD₇₅₀ with different light intensities on day two.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	0.0802	10.1755	<.0001
Algae type	1	0.0006	0.7928	0.3821
Light intensity	5	0.0547	15.2699	<.0001
Algae type* Light intensity	5	0.0249	6.9577	0.0004
Error	24	0.0609		
Total	35	0.0974		

Table 10. Analysis of variance (ANOVA) table of OD₇₅₀ with different light intensities on day three.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	0.3558	11.5330	<.0001
Algae type	1	0.1126	40.1300	<.0001
Light intensity	5	0.0712	5.0756	0.0026
Algae type* Light intensity	5	0.1721	12.2711	<.0001
Error	24	0.0673		
Total	35	0.4232		

Table 11. Analysis of variance (ANOVA) table of OD₇₅₀ with different light intensities on day four.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	1.1396	48.5503	<.0001
Algae type	1	0.4306	201.7673	<.0001
Light intensity	5	0.2966	27.7953	<.0001
Algae type* Light intensity	5	0.4125	38.6618	<.0001
Error	24	0.0512		
Total	35	1.1908		

Table 12. Analysis of variance (ANOVA) table of ETR with different light intensities on day two.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	3836.0716	15.3984	<.0001
Algae type	1	2437.3574	107.6218	<.0001
Light intensity	5	240.0179	2.1196	0.0978
Algae type* Light intensity	5	1158.6963	10.2325	<.0001
Error	24	543.5385		
Total	35	4379.6102		

Table 13. Analysis of variance (ANOVA) table of ETR with different light intensities on day three.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	4636.9867	14.6571	<.0001
Algae type	1	3741.3937	130.0882	<.0001
Light intensity	5	362.7087	2.5223	0.0570
Algae type* Light intensity	5	532.8843	3.7057	0.0126
Error	24	690.2505		
Total	35	5327.2372		

Table 14. Analysis of variance (ANOVA) table of ETR with different light intensities on day four.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	10203.237	68.6915	<.0001
Algae type	1	7199.7261	533.1801	<.0001
Light intensity	5	1339.7153	19.8427	<.0001
Algae type* Light intensity	5	1663.7960	24.6427	<.0001
Error	24	324.081		
Total	35	10527.318		

Figure 13. Growth of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* at different light intensities (from 100-600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$).

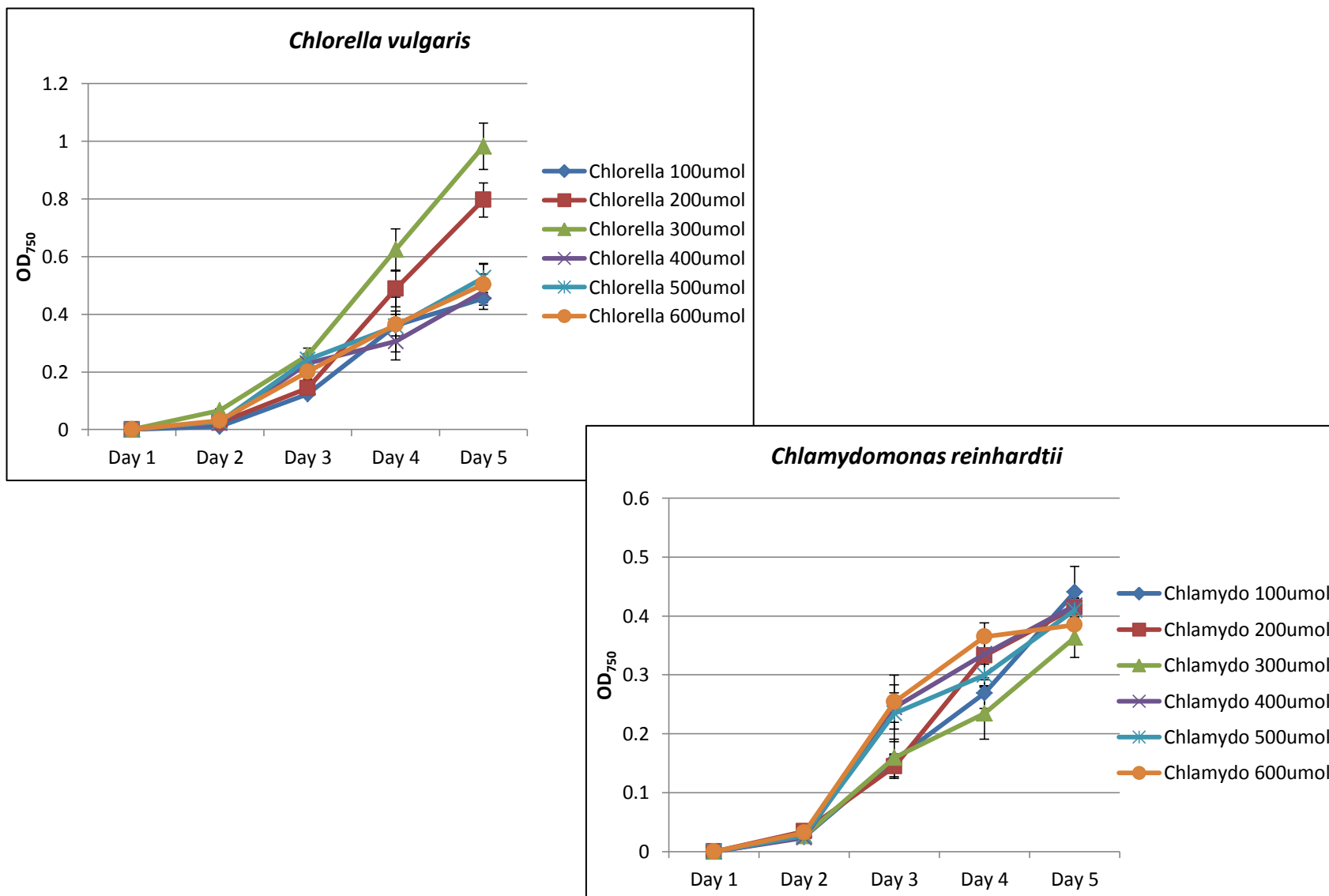


Figure 14. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative electron transport rates (ETR) at different light intensities (from 100-600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$) measured on day two, three and four.

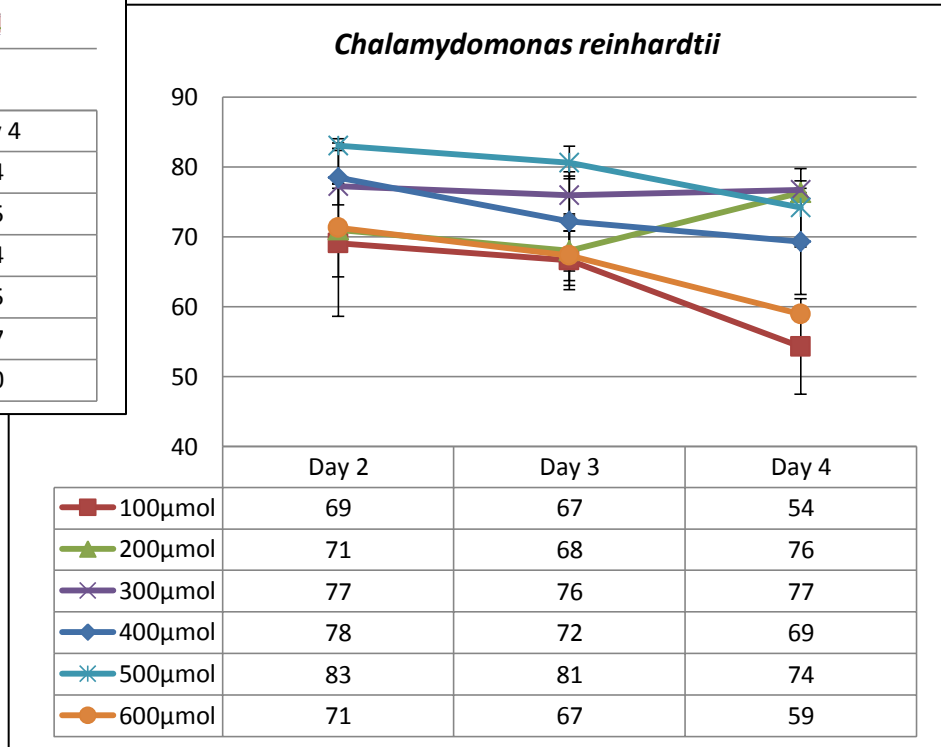
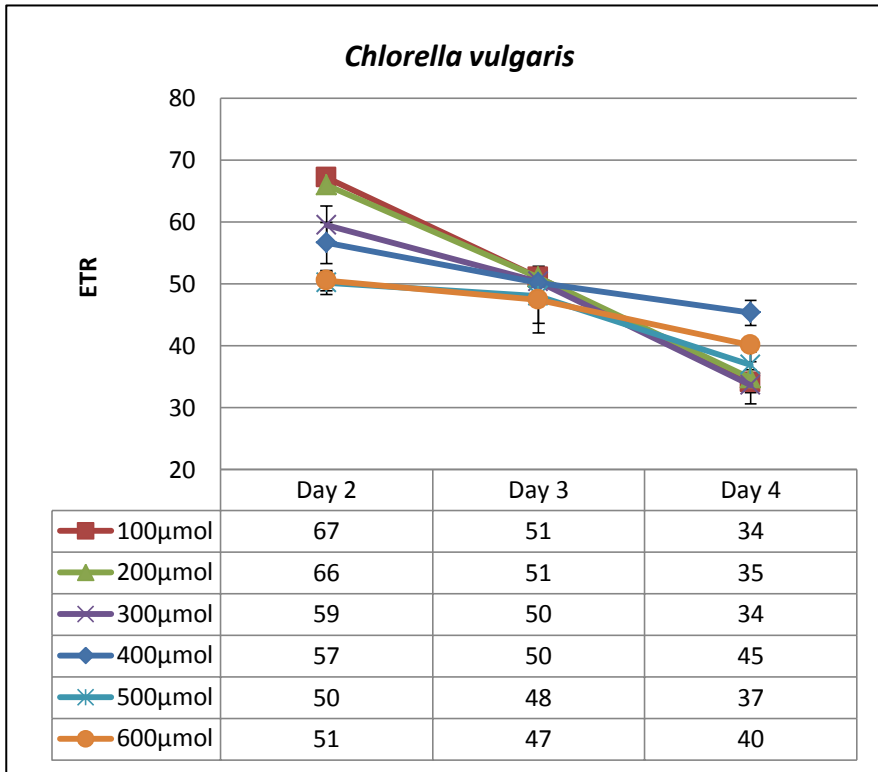


Figure 15. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative ETR (Electron Transport Rate) as a function of PAR (Photosynthetically Active Radiation) at light intensities of 100-300 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ measured on day two, three and four.

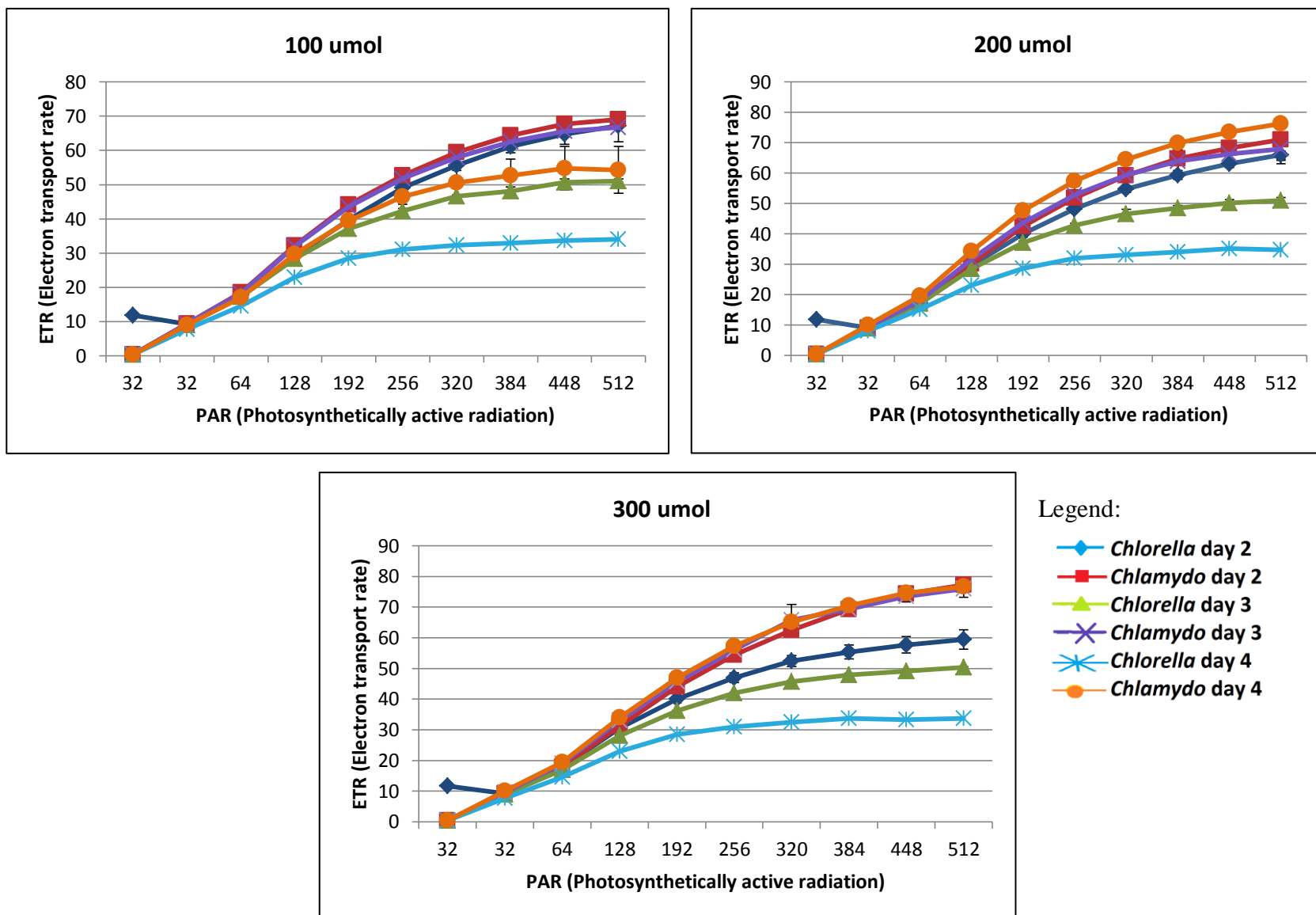


Figure 16. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative ETR (Electron Transport Rate) as a function of PAR (Photosynthetically Active Radiation) at light intensities of 400-600 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$ measured on day two, three and four.

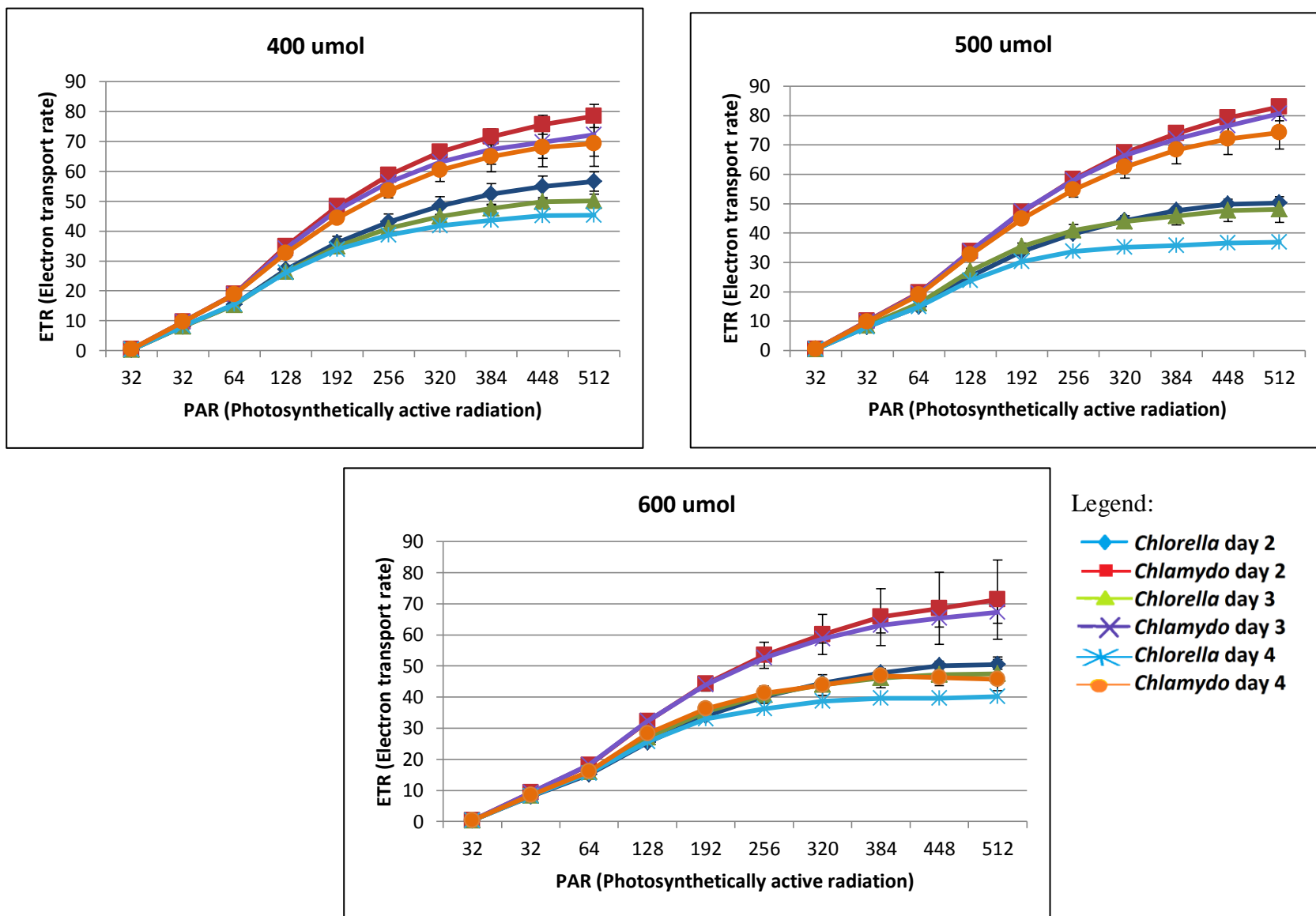


Figure 17. Growth of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* in ambient air (0.038%), 3% CO₂, 6% CO₂, and 12% CO₂.

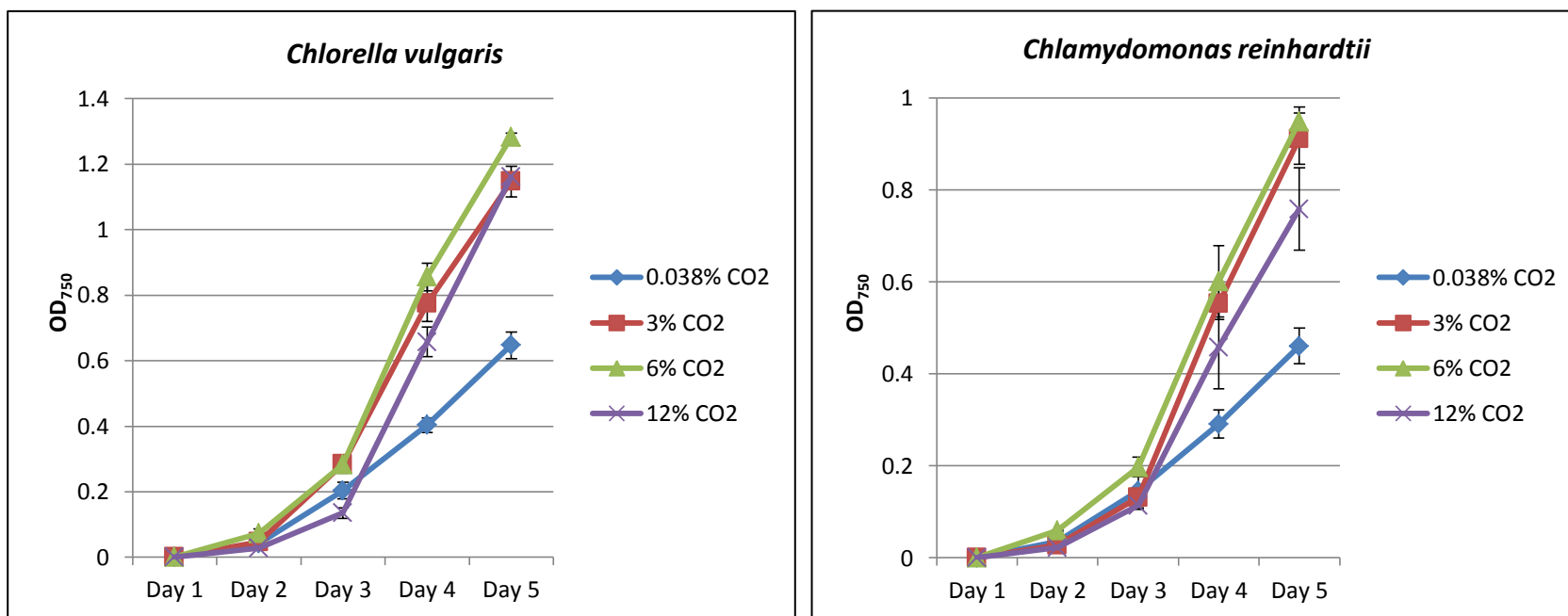


Table 15. Analysis of variance (ANOVA) table of OD₇₅₀ with different CO₂ concentrations on day two.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	0.0955	27.1731	<.0001
Algae type	1	0.0376	72.9149	<.0001
% CO ₂	3	0.0441	29.2650	<.0001
Algae type* % CO ₂	3	0.0138	9.1673	0.0009
Error	16	0.0080		
Total	23	0.1035		

Table 16. Analysis of variance (ANOVA) table of OD₇₅₀ with different CO₂ concentrations on day three.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	0.7515	36.0975	<.0001
Algae type	1	0.2322	79.0914	<.0001
% CO ₂	3	0.5027	56.3368	<.0001
Algae type* % CO ₂	3	0.0166	1.8604	0.1770
Error	16	0.0476		
Total	23	0.7991		

Table 17. Analysis of variance (ANOVA) table of OD₇₅₀ with different CO₂ concentrations on day four.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	1.6581	107.5106	<.0001
Algae type	1	0.5017	227.7067	<.0001
% CO ₂	3	1.1142	168.5734	<.0001
Algae type* % CO ₂	3	0.0421	6.3826	0.0047
Error	16	0.0353		
Total	23	1.6934		

Table 18. Analysis of variance (ANOVA) table of ETR with different CO₂ concentrations on day two.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	5889.1250	70.7829	<.0001
Algae type	1	3981.4656	334.9797	<.0001
% CO ₂	3	894.2447	25.0790	<.0001
Algae type* % CO ₂	3	1013.4147	28.4211	<.0001
Error	16	190.1711		
Total	23	6097.2961		

Table 19. Analysis of variance (ANOVA) table of ETR with different CO₂ concentrations on day three.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	6974.2753	17.7621	<.0001
Algae type	1	4291.9627	76.5155	<.0001
% CO ₂	3	1495.4897	8.8870	0.0011
Algae type* % CO ₂	3	1186.8230	7.0527	0.0031
Error	16	897.4841		
Total	23	7871.7594		

Table 20. Analysis of variance (ANOVA) table of ETR with different CO₂ concentrations on day four.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	7	6088.5057	869.7870	<.0001
Algae type	1	3629.7430	291.2247	<.0001
% CO ₂	3	2184.9535	58.4350	<.0001
Algae type* % CO ₂	3	273.8092	7.3228	0.0026
Error	16	199.4195		
Total	23	6287.9252		

Figure 18. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative electron transport rates (ETR) in ambient air (0.038%), 3% CO₂, 6% CO₂, and 12% CO₂ measured on day two, three and four.

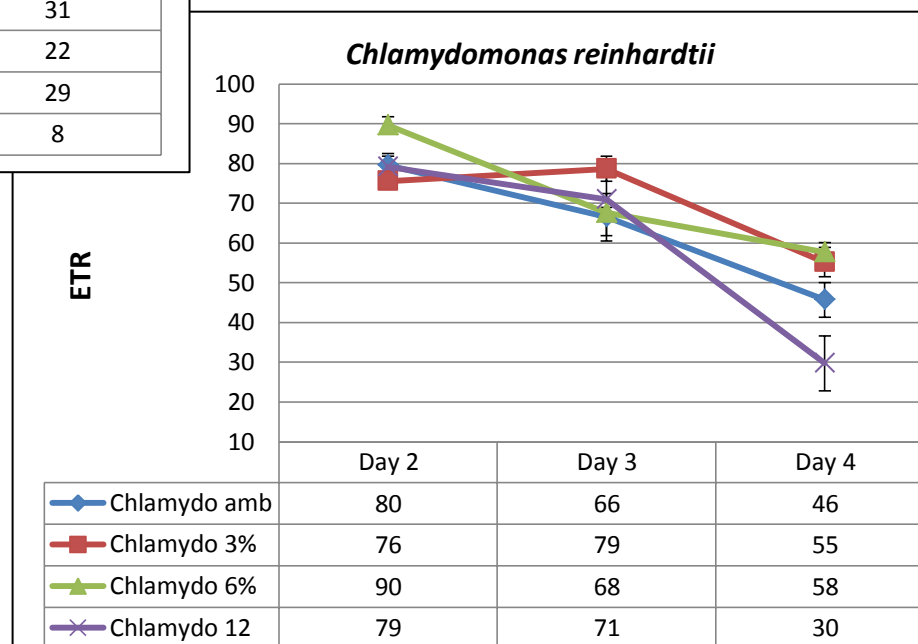
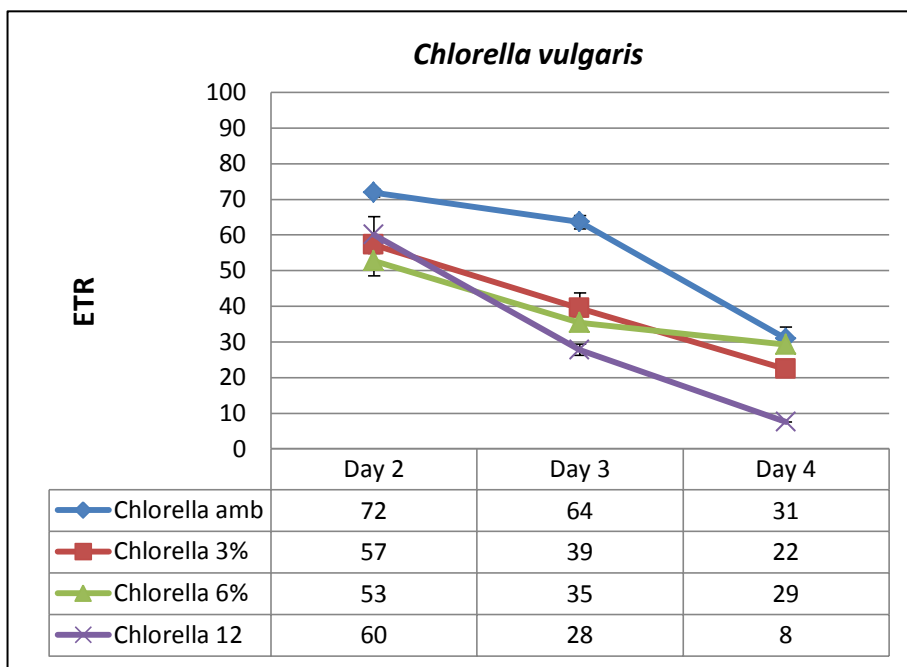
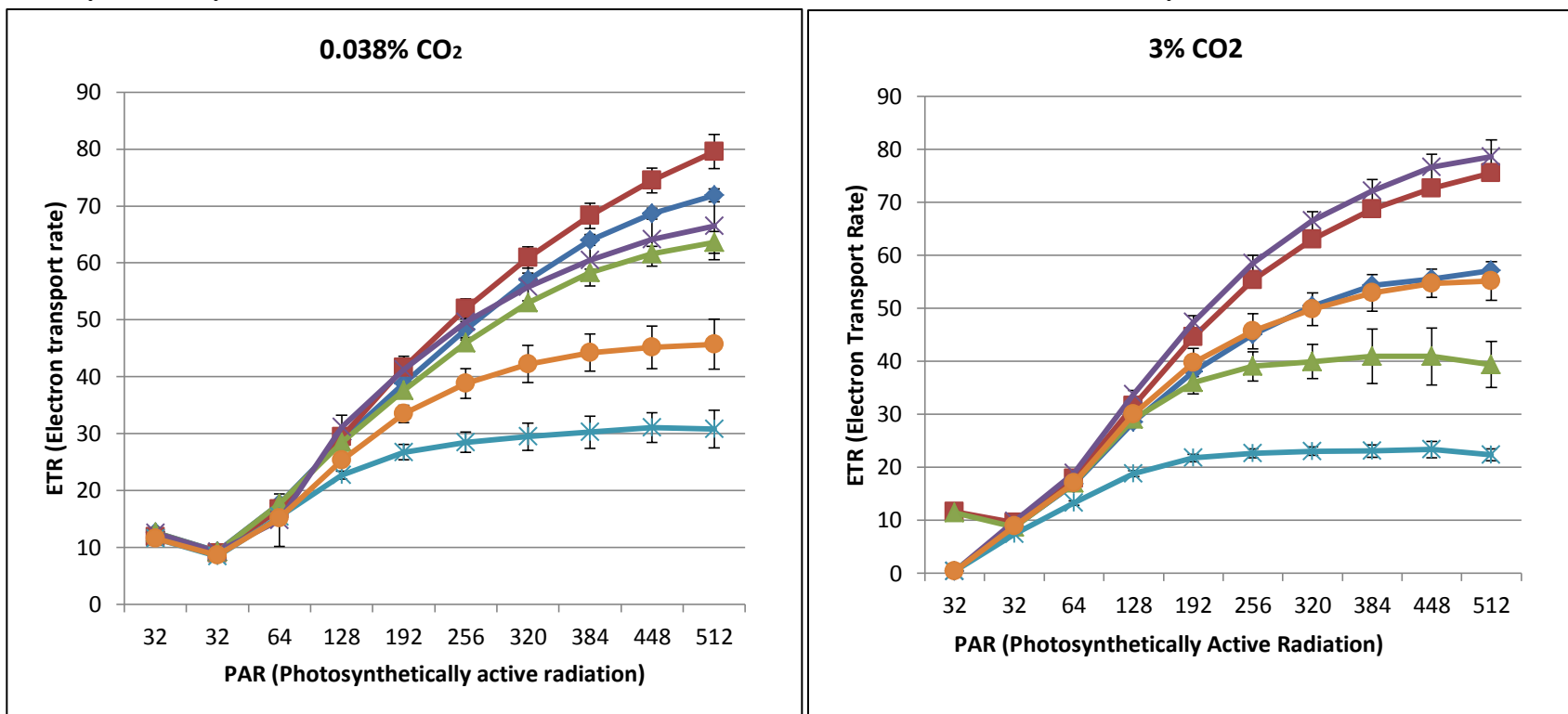


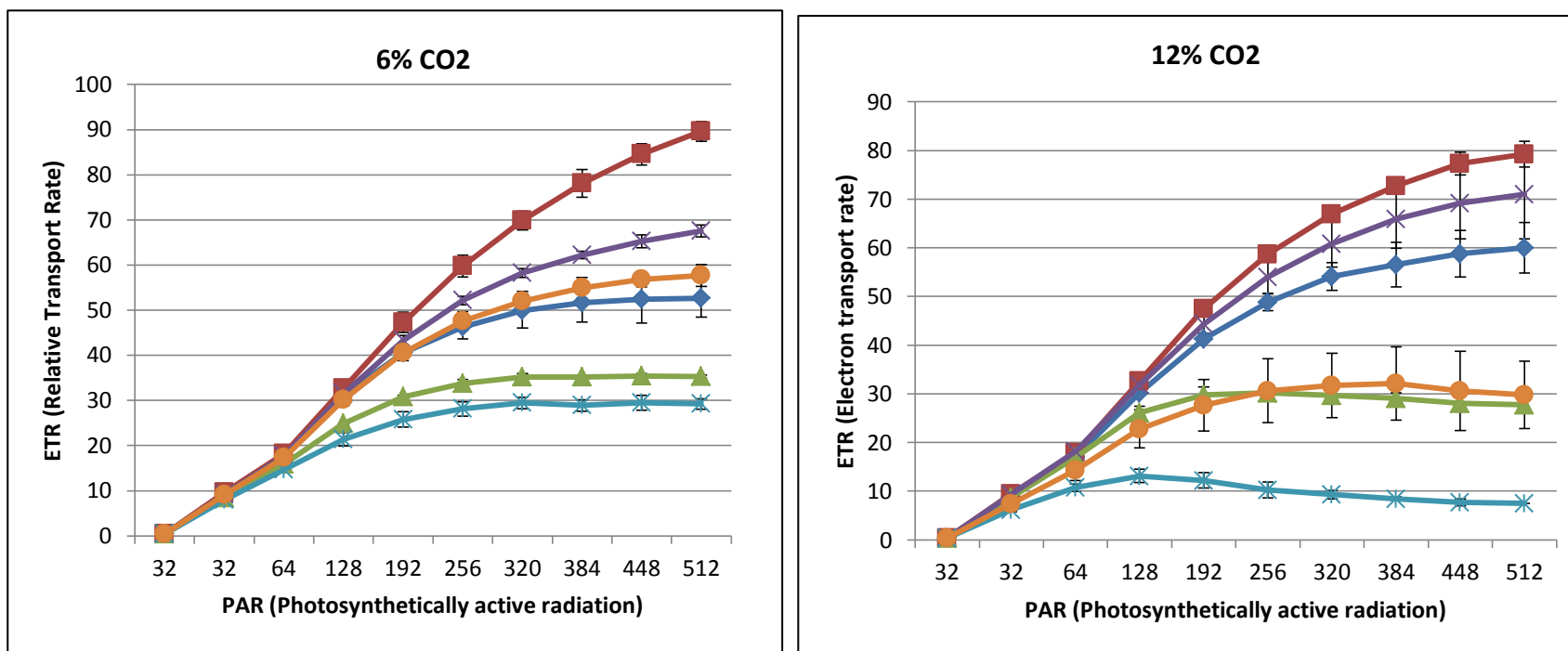
Figure 19. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative electron transport rates (ETR) as a function of PAR (Photosynthetically Active Radiation) at ambient CO₂ (0.038% CO₂) and 3% CO₂ measured on day two, three and four.



Legend:

- ◆ *Chlorella* day 2
- *Chlamydo* day 2
- ▲ *Chlorella* day 3
- ✖ *Chlamydo* day 3
- ✱ *Chlorella* day 4
- *Chlamydo* day 4

Figure 20. *Chlorella vulgaris* and *Chlamydomonas reinhardtii* relative electron transport rates (ETR) as a function of PAR (Photosynthetically Active Radiation) at 6% CO₂ and 12% CO₂ measured on day two, three and four.



Legend:

- ◆ *Chlorella* day 2
- *Chlamydo* day 2
- ▲ *Chlorella* day 3
- ✱ *Chlamydo* day 3
- ✱ *Chlorella* day 4
- *Chlamydo* day 4

Table 21. Analysis of variance (ANOVA table) of the total lipid content of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* grown at different CO₂ concentrations and using two types of extractions.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	137.3847	8.7883	0.0004
Algae type	1	32.7834	23.0683	0.0004
Extraction type	1	1.3680	0.9626	0.3459
Algae type*Extraction type	1	0.0294	0.0207	0.8880
% CO ₂	2	87.3055	30.7165	<.0001
Algae type*% CO ₂	2	10.2033	3.5898	0.0600
Extraction type*% CO ₂	2	2.7361	0.9626	0.4095
Algae type*Extraction type*% CO ₂	2	2.9589	1.0410	0.3829
Error	12	17.0538		
Total	23	154.4385		

Table 22. Analysis of variance (ANOVA table) of the total fatty acid content of *Chlorella vulgaris* and *Chlamydomonas reinhardtii* grown at different CO₂ concentrations and using two types of extractions.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	11	89.3821	6.4777	0.0016
Algae type	1	22.6095	18.0240	0.0011
Extraction type	1	0.0077	0.0062	0.9388
Algae type*Extraction type	1	0.2952	0.2353	0.6363
% CO ₂	2	56.1538	22.3826	<.0001
Algae type*% CO ₂	2	7.3585	2.9330	0.0918
Extraction type*% CO ₂	2	1.3754	0.5482	0.5918
Algae type*Extraction type*% CO ₂	2	1.5821	0.6306	0.5490
Error	12	15.0529		
Total	23	104.4350		

Table 23. Lipid content, total fat content and fatty acid content of *C. vulgaris* and *C. reinhardtii* (g/ 100g biomass) grown at ambient CO₂ concentration, 6 % CO₂ and 12% CO₂ using 8000M Mixer/Mill and Soxtec extraction system.

Algae type	Extraction type	% CO ₂	Lipid content	Total fat content	14:1	16:0	16:1	18:0	18:1n9	18:1n9	18:1n7	18:2	ω18:3	DHA
Chlorella	Mixer/Mill	0.038	13.1250 ^{abc}	9.0739 ^{abc}	0.1033 ^b	1.4058 ^{bcd}	0.3511 ^{ab}	0.1146 ^c	0.8606 ^a	0.7249 ^{ab}	0.3049	1.5657 ^{abcd}	1.6561 ^{ab}	0.0120 ^{cd}
Chlorella	Mixer/Mill	6	14.4350 ^{ab}	10.2462 ^{ab}	0.0960 ^b	1.8871 ^{abcd}	0.3603 ^{ab}	0.1609 ^{abc}	0.8307 ^a	0.8391 ^{ab}	0.3686	2.0463 ^{abcd}	1.4047 ^{abc}	0.0150 ^{bcd}
Chlorella	Mixer/Mill	12	17.7875 ^a	12.8995 ^a	0.1951 ^a	2.1841 ^{abcd}	0.4020 ^a	0.2335 ^{abc}	0.7359 ^{ab}	1.2059 ^{ab}	0.3364	2.7603 ^a	1.3134 ^{abc}	0.0258 ^a
Chlorella	Soxtec	0.038	12.8525 ^{bc}	9.5864 ^{abc}	0.0743 ^b	1.8301 ^{abcd}	0.3624 ^{ab}	0.1644 ^{abc}	0.9346 ^a	0.6869 ^{ab}	0.3015	1.5496 ^{abcd}	1.7965 ^a	0.0163 ^{abcd}
Chlorella	Soxtec	6	13.6700 ^{ab}	10.1419 ^{abc}	0.0892 ^b	2.1224 ^{abcd}	0.3814 ^{ab}	0.2058 ^{abc}	0.7279 ^{ab}	0.9333 ^{ab}	0.3139	1.9791 ^{abcd}	1.2373 ^{abcd}	0.0152 ^{bcd}
Chlorella	Soxtec	12	17.6025 ^a	11.7184 ^a	0.1219 ^b	2.2405 ^{abcd}	0.3838 ^{ab}	0.2442 ^{abc}	0.6764 ^{ab}	1.1129 ^{ab}	0.3062	2.5950 ^{ab}	1.1823 ^{abcd}	0.0223 ^{ab}
Chlamydo	Mixer/Mill	0.038	10.9600 ^{bc}	6.2355 ^{bc}	0.0923 ^b	1.2050 ^d	0.1262 ^d	0.1021 ^c	0.1806 ^c	0.2436 ^b	0.3063	0.6847 ^d	1.0194 ^{bcd}	0.0086 ^d
Chlamydo	Mixer/Mill	6	13.5975 ^{ab}	8.9830 ^{abc}	0.1144 ^b	1.6551 ^{abcd}	0.3100 ^{abc}	0.1486 ^{bc}	0.4477 ^{bc}	0.6724 ^{ab}	0.3400	1.3701 ^{bcd}	1.2647 ^{abcd}	0.0135 ^{bcd}
Chlamydo	Mixer/Mill	12	13.9875 ^{ab}	10.5122 ^{ab}	0.0610 ^b	2.6086 ^a	0.2070 ^{bcd}	0.3201 ^a	0.3899 ^{bc}	1.9415 ^a	0.2393	2.3252 ^{abc}	0.8171 ^{cd}	0.0189 ^{abc}
Chlamydo	Soxtec	0.038	8.5700 ^c	5.7432 ^c	0.0532 ^b	1.3424 ^{cd}	0.1402 ^{cd}	0.1482 ^{bc}	0.2470 ^c	0.5943 ^{ab}	0.2493	0.9909 ^{cd}	0.6367 ^d	0.0110 ^{cd}
Chlamydo	Soxtec	6	13.5200 ^{ab}	10.1397 ^{abc}	0.0522 ^b	2.4732 ^{ab}	0.2069 ^{bcd}	0.3099 ^a	0.4056 ^{bc}	2.0375 ^a	0.0000	2.2671 ^{abc}	0.8688 ^{cd}	0.0193 ^{abc}
Chlamydo	Soxtec	12	14.8125 ^{ab}	10.4056 ^{ab}	0.0817 ^b	2.4037 ^{abc}	0.2503 ^{abcd}	0.2888 ^{ab}	0.4433 ^{bc}	1.5667 ^{abc}	0.2114	2.0269 ^{abcd}	1.0479 ^{bcd}	0.0186 ^{abc}

Fig.21. Major fatty acid composition of *C. vulgaris* grown using different CO₂ concentrations.

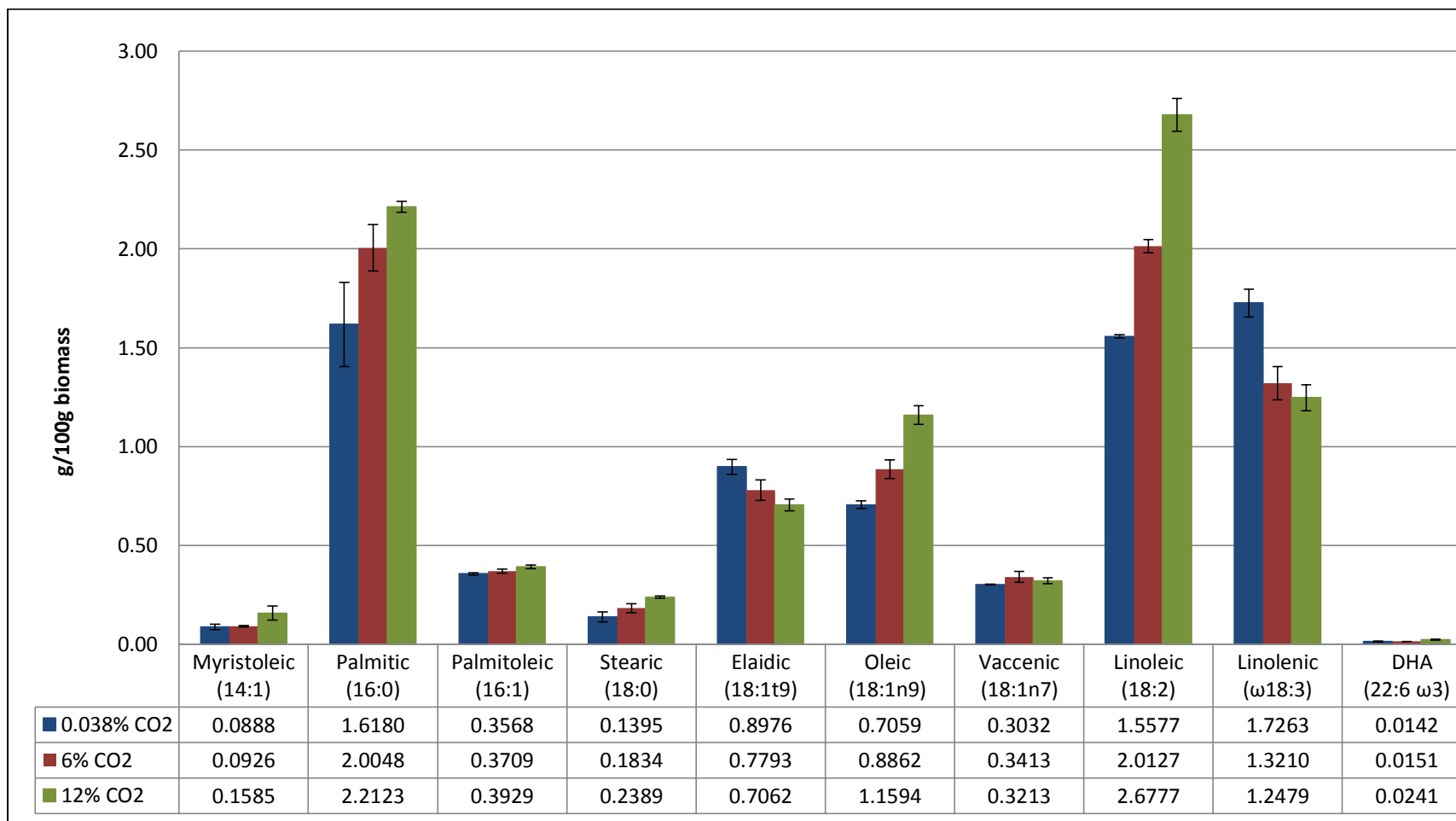
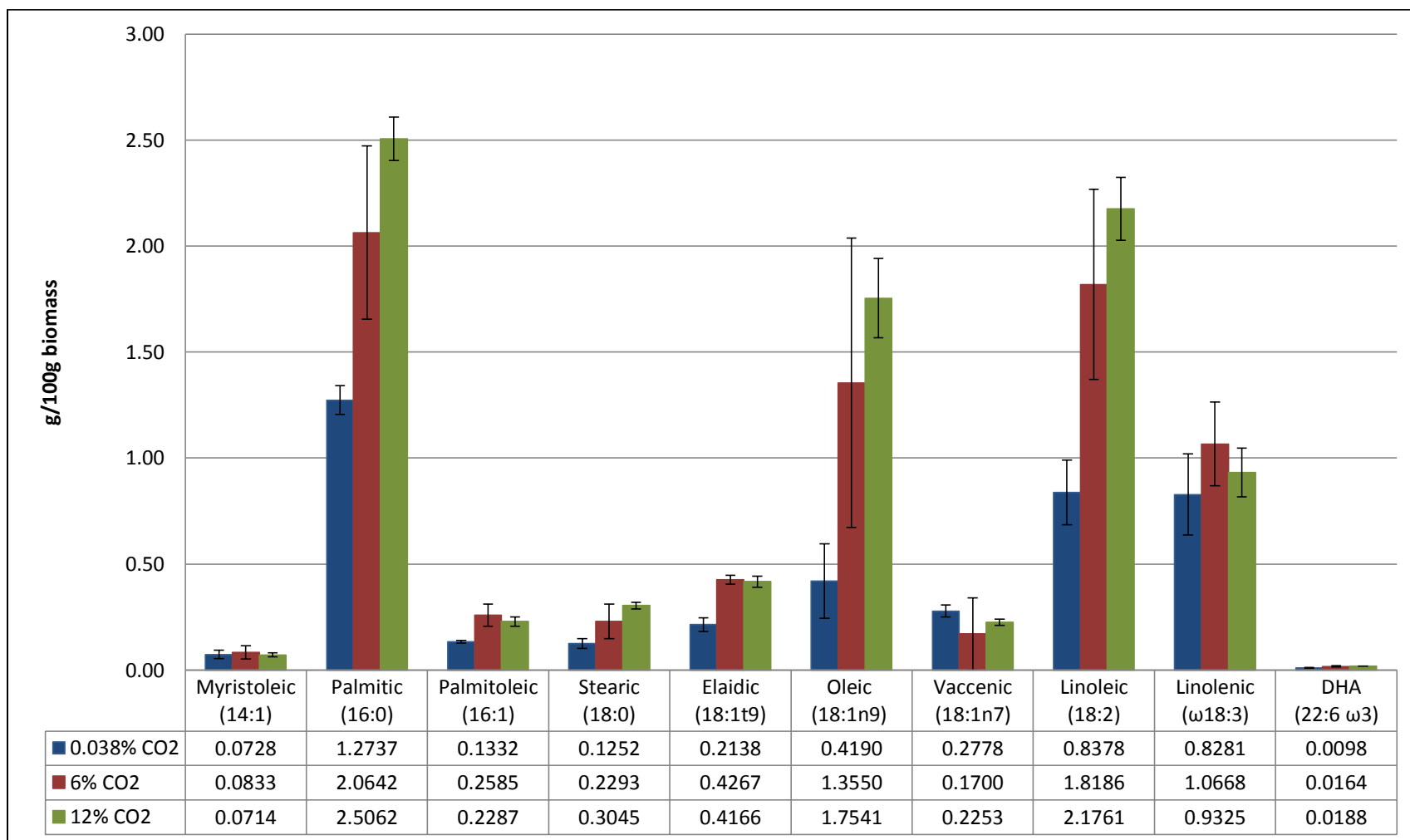


Fig.22. Major fatty acid composition of *C. reinhardtii* grown using different CO₂ concentrations.



CHAPTER 2

Comparison of CO₂ uptake by *Chlorella vulgaris* and *Chlamydomonas reinhardtii* in liquid culture and immobilized algal beads

2.1 INTRODUCTION

One of the today's biggest environmental challenges is emission of greenhouse gases (GHG) such as CO₂ from combustion of fuels (Melillo *et al.*, 1993). Common remediation approaches to managing the levels of CO₂ released into the atmosphere include 1) increased efficiency of energy conversion, 2) use of lower carbon energy sources, and 3) various forms of carbon sequestration (Keffer and Kleinheinz, 2002).

Microalgae can fix carbon dioxide from the atmosphere using open-culture systems (lakes or ponds) or closed-culture systems called photo-bioreactors. Open-culture systems are less expensive to build and operate, are more durable than bioreactors and have a large production capacity when compared with closed systems. Photo-bioreactors are advantageous because they are flexible systems that can be optimized according to the biological and physiological characteristics of the algal species being cultivated which allows cultivation of algal species that cannot be grown reliably in open ponds. Mitigation techniques using membrane-type photo-bioreactors have high CO₂ removal efficiency (Cheng *et al.*, 2006; Fan *et al.*, 2008). However, membrane-type photo-bioreactors have operational problems, which include membrane fouling and high design and operational costs (Cheng *et al.*, 2006). Therefore, tubular and bubble column-type photo-bioreactors are commonly applied for use in algal photoreactors (Chiu *et al.*, 2009; de Morais and Costa, 2007a; Fan *et al.*, 2008).

Biological CO₂ sequestration (accomplished by plants and photosynthetic microorganisms) is a promising strategy since the biomass produced can be further utilized for value-added products (de Morais and Costa, 2007a). Sequestration strategies include deep ocean injection, below ground storage of CO₂ in geological formations such as in oil and gas fields or saline aquifers (Stevens, 2000; Nordbotten *et al.*, 2005), and biological sequestration (de Morais and Costa, 2007a,b,c). The potential of carbon capture by green plants has been estimated to be 3-6% of fossil fuel emissions (Skjanes *et al.*, 2007). Microalgae have the ability to fix CO₂ during photosynthesis with efficiency 10 to 50 times greater than that of terrestrial plants (Li *et al.*, 2008; Usui and Ikenouchi, 1997). According to Sahoo (2010), one ton of algae can fix 0.36

tons of carbon, 0.6 tons of nitrogen and 0.008 tons of phosphorus. Microalgae and cyanobacteria species used for CO₂ mitigation include *Botryococcus braunii* (Yoo *et al.*, 2010), *Chlorella vulgaris* (Chen *et al.*, 2010; Cheng *et al.*, 2006), *Chlorella kessleri* (de Morais and Costa, 2007b), *Chlorocuccum littorale* (Ota *et al.*, 2009), *Scenedesmus sp.* (de Morais and Costa, 2007a; Ho *et al.*, 2010), *Chlamydomonas reinhardtii* (Packer, 2009) and *Spirulina sp.* (de Morais and Costa, 2007a,b,c). Microalgal species that grow well under the natural day–night cycle are suitable for large scale outdoor cultivation systems (Stewart and Hessami, 2005), and strains that can directly use the CO₂ in power-plant flue gas are preferred (Benemann, 1993; de Morais and Costa, 2007c; Maeda *et al.*, 1995). However, few *Chlorella sp.* and cyanobacteria species could grow well and achieve high CO₂ fixation ability (500–1800 mg L⁻¹ d⁻¹) under a relative high tolerance for temperature or CO₂ concentration (Maeda *et al.*, 1995; Ono *et al.*, 2007; Yue and Chen, 2005).

There are several advantages of using microalgae for bio-mitigation purposes. Firstly, microalgae have higher growth and CO₂ fixation rates compared with terrestrial plants (Chisti 2007; Li *et al.* 2008). Secondly, CO₂ can be completely recycled because it can be transformed to biofuel and other bioproducts (Li *et al.* 2008; Huntley and Redalje 2000). Thirdly, microalgae CO₂ fixation is more cost-effective and environmental sustainable when it is combined with other processes such as wastewater treatment (Wang *et al.*, 2008). However, to reduce CO₂ levels from industrial waste gas, microalgae have to be tolerant not only to high levels of CO₂, but they would also have to be able to grow in the presence of sulfur dioxides, nitrogen oxides, and volatile organic compounds (VOCs) which exist in the waste gases (Brown, 1996). Some researchers propose isolation of microalgae from lakes or ponds in the area of coal or oil fired thermoelectric power plants to obtain microalgae tolerant to the conditions prevalent in that particular area, in this way the adaptation of exotic strains to the new cultivation conditions existing in a power plant can be avoided (de Morais and Costa, 2007b).

There are several factors that play an important role in controlling the efficacy of CO₂ uptake by microalgae. High light intensity greater than saturation and high oxygen concentration can inhibit photosynthesis and CO₂ fixation (Pope, 1975; Jeon *et al.*, 2005). Basic growth nutrients are fundamental for maintaining the physiological integrity of the culture (Anderson, 2005). Temperature can be a decisive factor in the selection of the proper type of microalga with the purpose of CO₂ sequestration (Caron *et al.*, 1986). Sakai *et al.* (1995) isolated *Chlorella sp.* from hot springs in Japan able to grow at temperatures up to 42°C and more than 40% CO₂ in

air. Tolerance to both high temperature and high CO₂ content made this alga potentially suitable for bio-CO₂ mitigation from flue gas. Recent reports (Hsueh *et al.*, 2007; Zhang *et al.*, 2002) observed that the CO₂ fixation can be influenced by aeration rate, bubble size, and residence time of the bubble. In general, obtaining an extensive air/liquid interface area is the first vital step to improve the CO₂ mass transfer efficiency (Carvalho and Malcata, 2001; Ryu *et al.*, 2009).

Immobilized algae have become a new branch of biotechnology of rapidly growing importance (Hameed and Ebrahim, 2007). An immobilized cell is defined as a cell prevented from moving independently of its neighbors to all parts of the aqueous phase of the system, either by natural or artificial methods (Tampion and Tampion, 1987). The most frequent method used for algae immobilization is entrapment using alginate or carrageenan as polymers, where the cells are free within their compartments and the pores in the material allow substrates and products to diffuse to and from the cells (Mallick, 2002).

Studies on immobilized algae report varying results. Some studies found that the growth rates of immobilized cells are generally lower (Bailliez *et al.*, 1985; Robinson *et al.*, 1985) than those of the free cell cultures (Chevalier and de la Noue, 1985). Other studies suggest that there is no difference in oxygen evolution between free and immobilized *Chlorella* cells observed under a range of light intensities (Robinson *et al.*, 1985). Still other studies found that oxygen evolution was greater in the immobilized state, suggesting a fundamental change of metabolism (Bailliez *et al.*, 1988). Leon and Galvan (1995) studied the production of glycerol in *Chlamydomonas reinhardtii* cells immobilized in Ca-alginate and found that the immobilized cells showed a higher production rate in comparison to their free-living counterparts. Other studies demonstrated a decrease in productivity with immobilization. For example, keto-acid production by *Anacystis* and *Chlorella* was reduced by 70–90% (Wilkstrom *et al.*, 1982) following immobilization. The CO₂ fixation rates of algae immobilized in alginate beads are still largely unexplored. The present study is the first report of CO₂ uptake rates by *C. vulgaris* and *C. reinhardtii* alginate beads, using influent CO₂ concentration ranging from 3 to 9% CO₂.

The purpose of this study was to compare CO₂ fixation by *C. vulgaris* and *C. reinhardtii* grown in liquid cultures at ambient (0.038%), 3%, 6%, 9%, and 12% CO₂. The rates of CO₂ fixation on immobilized algal beads with a CO₂ influent of 3%, 6%, and 9% CO₂ were also identified.

2.2. MATERIALS AND METHODS

2.2.1. Consumption of CO₂ in liquid culture

2.2.1.1. Microalgal strain and culture conditions

Algae *Chlorella vulgaris* (UTEX 2714) and *Chlamydomonas reinhardtii* (UTEX 90) were grown at ambient (0.038%) CO₂, 3% CO₂, 6% CO₂, 9% CO₂, and 12% CO₂ for seven days using Wright's cryptophytes (WC-Gly) medium in 250 ml flasks, and stirred continuously. Temperature was maintained at 21°C and light intensity was maintained at 200 μmol m⁻² s⁻¹ with a photoperiod of 12 h light/12 h dark. The light energy for the photosynthetic fixation of CO₂ by the algae was provided by two GE grow lights source (General Electric, Fairfield, CT, USA).

2.2.1.2. Assay system and CO₂ measurements by gas chromatography

Cells were harvested by centrifugation and the final absorbance was adjusted to OD₇₅₀=0.050 (See Appendix 1). Three ml of the algal suspensions were added to 100 ml of Wright's Cryptophytes medium in five replicate 160 ml serum bottles and the head space was adjusted to 12% CO₂ (Figure 23). Each bottle corresponded to a specific time (0, 9, 21, 26, and 34 hrs. after inoculation). Another five bottles were not inoculated and used as controls. The bottles were kept under light on four stirrer plates (Cimarec, Thermo Scientific, USA). The experiment was run in duplicate. Three milliliters of gas were sampled using a 5 cc syringe (Becton Dickinson, NJ, USA) at each time point from each bottle, and the amount of CO₂ consumption of gas was monitored by gas chromatography with a Carle AGC Series 100 gas chromatograph (Hach Co., Loveland, CO, USA) connected to a BD40 chart recorder (Kipp & Zonen, Inc., Bohemia, NY, USA). From the same samples, the absorbance was measured at 750 nm using Spectronic 20 Genesys spectrophotometer (Spectronic Instruments, NY).

2.2.2. Consumption of CO₂ flowing through immobilized algal beads

2.2.2.1. Microalgal strain and culture conditions

C. vulgaris (UTEX 2714) and *C. reinhardtii* (UTEX 90) were grown in 250 ml flasks at ambient (0.038%) CO₂ using WC-Gly medium, 21°C temperature, and 200 μmol m⁻²s⁻¹ light intensity and stirred continuously.

Immobilized algal beads were prepared by pouring a concentrated suspension of algal cells (500 mL, OD₇₅₀=2.200, See Appendix 1) into an equal volume of 2.5% sodium alginate solution (Mallick, 2002) (Figure 24). The algae/alginate suspension was mixed for 2-3 hours using a stirrer until completely homogenized. Next, the suspension was drawn into a syringe and dripped slowly from the syringe tip into 100mM CaCl₂ solution to form the spherical beads. The beads were left overnight to cross-link the alginate molecules with calcium ions, trapping the cells in a matrix of calcium alginate. Next day, the beads were separated from the resulting sodium chloride solution and transferred into two 100 cm Tygon tubes (Thermo Fisher Scientific, Pittsburg, PA, USA) (ID=2.5 cm). For the control columns, beads were prepared in the same manner, without algal cells suspension.

2.2.2.2. Culture system and CO₂ measurements by gas chromatography

Gas in the range of 3% – 9% CO₂ was obtained by mixing compressed air with CO₂ from two tanks (Airgas Mid America, West Virginia, USA) in an stainless steel mixing chamber (20 cm length and an 4 cm internal diameter) (Figure 3). Gas flow through the columns was adjusted to 10-13 mL min⁻¹ for 3% and 6% CO₂ and 18-22 mL min⁻¹ for 9% CO₂. Influent (3, 6, and 9%) and effluent CO₂ concentrations were measured by gas chromatography with a Carle AGC Series 100 gas chromatograph (Hach Co., Loveland, CO, USA) connected to a BD40 chart recorder (Kipp & Zonen, Inc., Bohemia, NY, USA), over a five day period.

2.3. RESULTS

2.3.1. Consumption of CO₂ in liquid culture

ANOVA analysis showed that both *C. vulgaris* and *C. reinhardtii* had the same maximum uptake rate, but the CO₂ concentration used to grow the cells had a significant impact on uptake (Table 25). Overall, cells grown at higher CO₂ concentration exhibited lower CO₂ uptake compared with cells grown at ambient CO₂ concentration. The highest CO₂ fixation rate was 278 mg CO₂ L⁻¹ day⁻¹ in case of *C. vulgaris* grown at ambient CO₂ and the lowest CO₂ fixation rate was 117 mg CO₂ L⁻¹ day⁻¹ displayed by *C. reinhardtii* grown at 12% CO₂ (Table 24).

The OD₇₅₀ also was measured (Fig. 27). Overall *C. vulgaris* grew to higher densities compared with *C. reinhardtii*, and cells grown at lower CO₂ concentration grew better than cells grown at higher CO₂ concentration (Table 25). *C. vulgaris* had higher growth than *C. reinhardtii* at all CO₂ concentrations tested.

2.3.2. Consumption of CO₂ flowing through immobilized algal beads

The percentage of CO₂ reduction by *C. vulgaris* and *C. reinhardtii* was measured as the difference of influent CO₂ and effluent CO₂ measured at both sample ports (See Fig. 25, port D and port F), for each treatment (3%, 6% and 9% CO₂ influent). The difference between influent CO₂ and effluent CO₂ measured for control columns also was taken into account. The percentage of CO₂ reduction was examined at day one versus day five. On day one, immobilized *C. vulgaris* exhibited significantly higher capacity for CO₂ reduction, compared with *C. reinhardtii*. *C. vulgaris* sequestered 34%, 13% and 5% of CO₂ from influent atmospherics containing 3%, 6% and 9% CO₂, respectively. In comparison, *C. reinhardtii* sequestered 26%, 12% and 4%. Both strains showed higher fixation efficiency on day one. Fixation efficiency declined over time (Table 26, Figure 28, 29, and 32).

The interaction between algae type and influent CO₂ concentration was not significant (Table 26 and 27). However, when rates were estimated without considering controls, higher rates of 6% CO₂ consumption were apparent for both *C. vulgaris* and *C. reinhardtii*. On day one,

rates estimates varied from 35 to 53 mg CO₂ L⁻¹ day⁻¹ (Table 30). These rates are considerably lower than those observed using liquid batch cultures.

2.4. DISCUSSION

2.4.1. Consumption of CO₂ in liquid culture

Algal cells grown at ambient CO₂ levels were able to fix CO₂ more rapidly than cells grown at higher CO₂ concentrations. Similar findings were reported by Chinnasamy *et al.* (2009). They demonstrated that CO₂ uptake by *C. vulgaris* decreased by 29% under elevated CO₂ (6% CO₂). They reported that algal cells grown at the ambient levels of CO₂ and higher temperature (40°C) exhibited a 23% growth rate of CO₂ uptake. Increase in temperature significantly enhances the process of photorespiration thus causing depletion of intracellular CO₂. To the contrary, DeLucia *et al.* (1985) found that the rate of photosynthesis initially significantly increased for a short period at high CO₂ concentrations followed by a gradual decrease.

Table 24 presents a comparison between the carbon fixation rates indicated in the literature and the results of this study. The differences could be due to different conditions of growth (different media, light intensity, CO₂ concentrations, and pH) and type of reactors (tubular, bubble column, membrane, and air-lift) used in the various studies. Badger and Price (1994) explained that the efficiency of dissolved inorganic carbon (DIC) utilization for photosynthesis in low-CO₂ grown cells was higher due to the activity of carbonic anhydrase (an enzyme that catalyzes the interconversion of CO₂ and HCO³⁻). Carbonic anhydrase activity in low-CO₂ grown cells was higher than that of cells grown in high-CO₂. Also, cells grown in low-CO₂ exhibited lower photorespiration, a lower CO₂ compensation point, and low O₂ inhibition of photosynthesis.

A reason for lower CO₂ consumption rates obtained in the present study might be due to difficulty of mixing all the batch culture on the stirrer plates, and unequal distribution of light. As presented in the previous chapter, light intensity can affect photosynthesis and algal growth. Mixing is important for better diffusion of carbon dioxide in the medium, and preventing sedimentation of algal cells. Algae can fix carbon dioxide from carbon dioxide and

hydrogencarbonate dissolved in their aqueous surroundings, but must overcome the low rate of carbon dioxide diffusion in water. The key enzyme that catalyzes the fixation of carbon dioxide, ribulose-1, 5-bisphosphate (RuBP) carboxylase/oxygenase (Rubisco) is localized in the chloroplast pyrenoid (center of carbon dioxide fixation within the chloroplast). CO₂ is the substrate required by Rubisco and not hydrogencarbonate, therefore the accumulated HCO³⁻ must be converted to CO₂ before inorganic carbon fixation takes place. The carbonic anhydrase catalyses the interconversion of hydrogencarbonate into carbon dioxide and results in a high local concentration of carbon dioxide, which Rubisco can use rapidly before the CO₂ has a chance to leak out of the cell (Moroney and Somachi, 1999).

Another method to improve the CO₂ fixation rate could be changing the light cycle from 12 h light/12 h dark to 24 h light/0 h dark. Jacob-Lopes *et al.* (2009) reported an increase rate of CO₂ fixation from 562 mg CO₂ L⁻¹ day⁻¹ to 1,440 mg CO₂ L⁻¹ day⁻¹ when such a change was implemented.

2.4.2. Consumption of CO₂ flowing through immobilized algal beads

The CO₂ reduction efficiency of alginate-immobilized *C. vulgaris* and *C. reinhardtii* decreased as the CO₂ influent was increased. One factor that confounds this result is the different flow rate at the different CO₂ concentrations tested. The column flows were 10-13 mL min⁻¹ for 3% and 6% CO₂ and 18-22 mL min⁻¹ for 9% CO₂. Unfortunately, this was unavoidable due to technical difficulties in achieving a stable 9% CO₂. Similarly, Doucha *et al.* (2005) reported a 10-50% decrease in CO₂ mitigation by *Chlorella* sp. in an outdoor open thin-layer photobioreactor due to increasing flue gas rate. Also, often it was not possible to maintain a constant CO₂ concentration for 24 hrs, therefore it was necessary to recalibrate the gas composition each morning. For future experiments a more stable flow rate and CO₂ concentrations will give a better idea of the CO₂ fixation efficiency of immobilized algal beads.

On day one *C. vulgaris* had a significant higher uptake compared with *C. reinhardtii*, but on day five there was not a significant difference.

Only one study has evaluated CO₂ capture by immobilized algae. Rooke *et al.* (2011) studied immobilized *C. vulgaris* and *Botryococcus braunii* cells in low sodium silica gel, and demonstrated growth inhibition due to competition for space. They also reported that the mesoporosity of the gels enabled diffusion of nutrients and gases, and promoted the light and

dark photosynthetic reactions. Furthermore, the activity and the viability of encapsulated cells were for at least 80 days. During this time, the efficiency of the photosynthetic bioreactor in terms of CO₂ remediation was monitored and suggested the capability for long term productivity of living gels. However, they noticed a decrease in photosynthetic yield. They assumed that this was an indicator of nutrient stress and suggested the need for improvement of gas diffusion, replenishment of nutrients, or cell acclimatization prior immobilization. Interestingly, also they noticed a differentiation in the durability of living gels in term of algal species. *C. vulgaris* cells were more viable and consistent to the matrix than *Botryococcus braunii*.

When was observed the experimental columns on day five, a general decrease in bead diameter. This might happen because air flowing through the columns dried out the beads, shrinking the beads and decreasing the space available to living algal cells. In future experiments it would be beneficial to flow the influent gas through sterile liquid media to maintain the alginate beads in a high humidity atmosphere.

Future studies could include the efficient removal of wastewater nitrogen and phosphorus by alginate-immobilized *C. vulgaris* and *C. reinhardtii*. Chevalier and de la Noue (1985) reported that immobilized *Scenedesmus* was capable of removing 90% of the ammonium (within four hours) and 100% of phosphate (within two hours) from a typical effluent, suggesting possible uses in the tertiary treatment of wastewaters. Similar findings were reported by Tam *et al.* (2000) who used *Chlorella vulgaris* cells immobilized in alginate beads for removing of ammonia and phosphate from wastewater. They also suggested that immobilized *Chlorella vulgaris* can be used as a secondary treatment process for domestic wastewater. Wang and Huang (2003) co-immobilized *Chlorella pyrenoidosa* and activated sludge for nitrate and phosphate removal. They reported 80% nitrate removal and 88% phosphate removal.

2.5. CONCLUSION

CO₂ fixation was explored using two approaches: 1) consumption of known quantities of CO₂ in sealed serum bottles, 2) consumption of CO₂ flowing through immobilized algal beads. In liquid culture, the cells grown at higher CO₂ concentration exhibited lower CO₂ uptake compared with algae grown at ambient CO₂ concentration. In the immobilized algae study, a significant decrease in CO₂ fixation from day one to day five and a decrease in CO₂ uptake with increasing the influent CO₂ concentration were evident. Both cases showed a similar trend in CO₂ fixation rate based on increasing CO₂ concentration level. The fixation rate decreased with increasing CO₂ concentration.

CO₂ consumption generally decreased over the five day experiment. The maximum CO₂ consumption rate observed using immobilized algae was only 20% of the maximum obtained in liquid culture. Clearly, additional work is needed to optimize and sustain CO₂ uptake in alginate beads.

2.6. REFERENCES

- Anderson R.A.** 2005. Algal Culturing Techniques. Elsevier Academic Press, Burlington.
- Badger MR, Price GD.** 1994. The role of carbonic anhydrase in photosynthesis. *Annu Rev Plant Physiol Plant Mol Biol.* **45**:369–392.
- Bailliez C, Largeau C, Casadevall E.** 1985 Growth and hydrocarbon production of *Botryococcus braunii* immobilized in calcium alginate gel. *Appl Microbiol Biotechnol.* **23**: 99–105.
- Bailliez C, Largeau C, Casadevall E, Yang LW, Berkaloff C.** 1988. Photosynthesis, growth and hydrogen production of *Botryococcus braunii* immobilized by entrapment and adsorption in polyurethane foams. *Appl Microbiol Biotechnol.* **29**: 141–147.
- Benemann J. R.** 1993. Utilization of carbon dioxide from fossil fuel - burning power plants with biological system. *Energy conversion and management.* **34**:999-1004.
- Brown LM.** 1996. Uptake of carbon dioxide from flue gas by microalgae. *Energy Convers. Mgmt.* **37**:1363-1367.
- Carvalho AP, Malcata FX.** 2001. Transfer of carbon dioxide within cultures of microalgae: plain bubbling versus hollow-fiber modules. *Biotechnol Prog.* **17**:265-72.
- Caron DA, Goldman JC, Dennet MR.** 1986. Effect of temperature on growth, respiration, and nutrient re-generation by an omnivorous microflagellate. *Applied and Environmental Microbiology.* **52**: 1340-1347.
- Chen CY, Yeh KL, Su HM, Lo YC, Chen WM, Chang JS.** 2010. Strategies to enhance cell growth and achieve high-level oil production of a *Chlorella vulgaris* isolate. *Biotechnol Prog,* **26**: 679-686.
- Cheng LH, Zhang L, Chen HL, Gao CJ.** 2006. Carbon dioxide removal from air by microalgae cultured in a membrane-photobioreactor. *Sep Purif Technol,* **50**: 324–329.
- Chevalier P, de la Noüe J.** 1985. Wastewater nutrient removal with microalgae immobilized in carrageenan. *Enzyme Microb Technol.* **7**: 621–624.
- Chinnasamy S, Ramakrishnan B, Bhatnagar A, KC Das.** 2009. Biomass production potential of a wastewater alga *Chlorella vulgaris* ARC 1 under elevated levels of CO₂ and temperature. *Int. J. Mol. Sci.* **10**: 518-532.
- Chiu SY, Kao CY, Chen CH, Kuan TC, Ong SC, Lin CS.** 2008. Reduction of CO₂ by a high-density culture of *Chlorella* sp in a semicontinuous photobioreactor. *Bioresour Technol,* **99**: 3389–3396.

- Chiu SY, Tsai MT, Kao CY, Ong SC, Lin CS.** 2009. The air-lift photobioreactors with flow patterning for high-density cultures of microalgae and carbon dioxide removal. *Eng Life Sci*, **9**: 254–260.
- Chisti Y.** 2007. Biodiesel from microalgae. *Biotechnol Adv.* **25**:294–306.
- Delucia EH, Sasek TW, Strain BR.** 1985. Photosynthetic inhibition after long-term exposure to elevated levels of atmospheric carbon dioxide. *Photosynthesis Research.* **7**:175-184.
- de Morais MG, Costa JAV.** 2007a. Biofixation of carbon dioxide by *Spirulina* sp. and *Scenedesmus obliquus* cultivated in a three stage serial tubular photobioreactor. *J Biotechnol* **129**:439–445.
- de Morais MG, Costa JAV.** 2007b. Carbon dioxide fixation by *Chlorella kessleri*, *C. vulgaris*, *Scenedesmus obliquus* and *Spirulina* sp. cultivated in flasks and vertical tubular photobioreactors. *Biotechnol Lett*, **29**: 1349–1352.
- de Morais MG, Costa JAV.** 2007c. Isolation and selection of microalgae from coal fired thermoelectric power plant for biofixation of carbon dioxide. *Energy Conversion and Management.* **48**: 2169-2173.
- Doucha J, Straka F, Livansky K.** 2005. Utilization of flue gas for cultivation of microalgae (*Chlorella* sp.) in an outdoor open thin-layer photobioreactor. *J Appl Phycol*, **17**: 403-412.
- Fan LH, Zhang YT, Zhang L, Chen HL.** 2008. Evaluation of a membrane-sparged helical tubular photobioreactor for carbon dioxide biofixation by *Chlorella vulgaris*. *J Membr Sci*, **325**: 336–345.
- Hanagata N, Takeuchi T, Fukuju Y, Barnes DJ, Karube I.** 1992. Tolerance of microalgae to high CO₂ and high temperature. *Phytochemistry.* **31**: 3345-3348.
- Hameed MSA, Ebrahim OH.** 2007. Review. Biotechnological potential uses of immobilized algae. *International journal of agriculture and biology.* **9**: 183–192.
- Hessen DO, Anderson TR.** 2008. Excess carbon in aquatic organisms and ecosystems: physiological, ecological, and evolutionary implications. *Limnology and Oceanography.* **53**: 1685-1696.
- Hirata S, Hayashitani M, Taya M, Tone S.** 1996a. Carbon dioxide fixation in batch culture of *Chlorella* sp. using a photobioreactor with a sunlight-collection device. *Journal of fermentation and bioengineering.* **81**: 470-472.
- Hirata S, Taya M, Tone S.** 1996b. Characterization of *Chlorella* cell cultures in batch and continuous operations under a photoautotrophic condition. *Journal of chemical engineering of Japan.* **29**: 953-959.

Huntley ME, Redalje DG. 2006. CO₂ mitigation and renewable oil from photosynthetic microbes: a new appraisal. *Mitigation and Adaptation Strategies for Global Change*. *Mitigation and Adaptation Strategies for Global Change* **12**: 573-608.

Hsueh HT, Chu H, Yu ST. 2007. A batch study on the bio-fixation of carbon dioxide in the absorbed solution from a chemical wet scrubber by hot spring and marine algae *Chemosphere*, **66**: 878–886.

Jacob-Lopes E, Scoparo C, Lacerda L, Franco T. 2009. Effect of light cycles (night/day) on CO₂ fixation and biomass production by microalgae in photobioreactors. *Chemical Engineering and Processing Process Intensification*. **48**: 306-310.

Jeon YC, Cho CW, Yu YS. 2005. Measurement of microalgal photosynthetic activity depending on light intensity and quality. *Biochemical Engineering Journal*. **27**: 127-131.

Keffer JE, Kleinheinz GT. 2002. Use of *Chlorella vulgaris* for CO₂ mitigation in a photobioreactor. *Journal of Industrial Microbiology and Biotechnology*. **29**: 275 – 280.

Leon R, Galvan F. 1995. Glycerol photoproduction by free and calcium-entrapped cells of *Chlamydomonas reinhardtii*. *J Biotechnol*. **42**: 61–67.

Li Y, Horsman M, Wu N, Lan CQ, Dubois-Calero N. 2008. Biofuels from microalgae. *Biotechnol. Prog.* **24**: 815-820.

Maeda K, Owada M, Kimura N, Omata K, Karube I. 1995. CO₂ fixation from the flue gas on coal-fired thermal power plant by microalgae. *Energy Convers. Mgmt.* **36**: 717-720.

Mallick N. 2002. Biotechnological potential of immobilized algae for wastewater N, P and metal removal: A review. *BioMetals*. **15**: 377–390.

Melillo JM, Mcguire AD, Kicklighter DW, Moore B, Vorosmarty CJ, AL Schloss. 1993. Global climate-change and terrestrial net primary production. *Nature*. **363**: 234–240.

Miyachi1 S, Iwasaki I, Shiraiwa Y. 2003. Historical perspective on microalgal and cyanobacterial acclimation to low- and extremely high-CO₂ conditions. *Photosynthesis Research*. **77**: 139–153.

Moroney JV, Somachi A. 1999. How do algae concentrate CO₂ to increase the efficiency of photosynthetic carbon fixation? *Plant Physiology*. **119**: 1 9-16.

Nordbotten JM, Celia MA, Bachu S. 2005. Injection and storage of CO₂ in deep saline aquifers: Analytical solution for CO₂ plume evolution during injection. *Transport in Porous Media*. **58**: 339-360.

Ono E, Cuello JL. 2003. Selection of optimal microalgae species for CO₂ sequestration. Proceedings 2nd annual conference on carbon sequestration. Alexandria, pp. 1-7.

Ono E, Cuello JL. 2007. Carbon dioxide mitigation using thermophilic cyanobacteria. *Biosys Eng*, **96**: 129–134.

Ota M, Kato Y, Watanabe H, Watanabe M, Sato Y, Smith RL. 2009. Fatty acid production from a highly CO₂ tolerant alga, *Chlorocuccum littorale*, in the presence of inorganic carbon and nitrate. *Bioresour Technol*, **100**: 5237–5242.

Packer M. 2009. Algal capture of carbon dioxide; biomass generation as a tool for greenhouse gas mitigation with reference to New Zealand energy strategy and policy. *Energy Policy*. **37**: 3428–3437.

Papazia A, Makridisb P, Divanachb P, Kotzabasisa K. 2008. Bioenergetic changes in the microalgal photosynthetic apparatus by extremely high CO₂ concentrations induce an intense biomass production. *Physiologia Plantarum*. **132**: 338-349.

Pope DH. 1975. Effects of light intensity, oxygen concentration, and carbon dioxide concentration on photosynthesis in algae. *Microbial Ecology*. **2**: 1-16.

Robinson, P.K., Dainty, A.L., Goulding, K.H., Simpkins, I., Trevan, M.D. 1985. Physiology of alginate-immobilized *Chlorella*. *Enzyme Microb Technol*. **7**: 212–216.

Rooke JC, Leonard A, Sarmento H, Meunier CF, Descyb J-P, Su B-L. 2011. Novel photosynthetic CO₂ bioconverter based on green algae entrapped in low-sodium silica gels. *J. Mater. Chem.*, **21**: 951–959.

Ryu HJ, Oh KK, Kim YS. 2009. Optimization of the influential factors for the improvement of CO₂ utilization efficiency and CO₂ mass transfer rate. *J Ind Eng Chem*, **15**: 471–475.

Sahoo D. 2010. Role of algae in carbon sequestration. In: Qasim, S.Z., and Goel, M., ed. CO₂ sequestration technologies for clean energy. Daya Publishing House, Dehli, pp. 111.

Sakai N, Sakamoto Y, Kishimoto N, Chihara M, Karube I. 1995. *Chlorella* strains from hot-springs tolerant to high-temperature and high CO₂. *Energy Convers Manage*, **36**: 693–696.

Skjanes K, Lindblad P, Muller J. 2007. BioCO₂—a multidisciplinary, biological approach using solar energy to capture CO₂ while producing H₂ and high value products. *Biomol Eng* **24**:405–413.

Stevens S. 2000. Sequestration of CO₂ in depleted oil & gas fields: Global capacity, costs and Barriers. In 5th International conference on greenhouse gas control technologies (GHGT-5), Cairns, Queensland, Australia.

- Stewart C, Hessami M-A.** 2005. A study of methods of carbon dioxide capture and sequestration — the sustainability of a photosynthetic bioreactor approach. *Energy Convers Manage*, pp. 403–420.
- Tam NFY, Wong YS.** 2000. Effect of immobilized microalgal bead concentrations on wastewater nutrient removal. *Environmental Pollution*. **107**:145-151.
- Tampion J, Tampion M.D.** 1987. *Immobilized cells. Principles and Applications*. Cambridge, UK. Cambridge University Press.
- Tsuzuki M, Gantar M, Aizawa K, Miyachi S.** 1986. Ultrastructure of *Dunaliella Tertiolecta* cells grown under low and high CO₂ Concentrations. *Plant Cell Physiology*. **27**: 737-739.
- Usui N, Ikenouchi M.** 1997. The Biological CO₂ Fixation and Utilization Project by RITE(1). *Energy Convers. Mgmt*. **38**: 487-492.
- Wang Y, Huang G.** 2003. Nitrate and phosphate removal by co-immobilized *Chlorella pyrenoidosa* and activated sludge at different pH. *Wat. Qual. Res. J. Canada*, **38**: 541–51.
- Wang B, Li Y, Wu N, Lan CQ.** 2008. CO₂ bio-mitigation using microalgae. *Appl Microbiol Biotechnol*. **79**:707–718.
- Yue L, Chen W.** 2005. Isolation and determination of cultural characteristics of a new highly CO₂ tolerant fresh water microalgae. *Energy Conversion and Management*, **46**: 1868-1876.
- Yoo C, Jun SY, Lee JY, Ahn CY, Oh HM.** 2010. Selection of microalgae for lipid production under high levels carbon dioxide. *Bioresour Technol*. **101**: S71–S74.
- Zhang K, Kurano N, Miyachi S.** 2002. Optimized aeration by carbon dioxide gas for microalgal production and mass transfer characterization in a vertical flat-plate photobioreactor. *Bioprocess Biosyst Eng*, **25**: 97–101.

Figure 23. Sealed serum bottles used for CO₂ consumption and growth measurements inoculated with *Chlorella vulgaris*.



Figure 24. Production of immobilized algal beads.

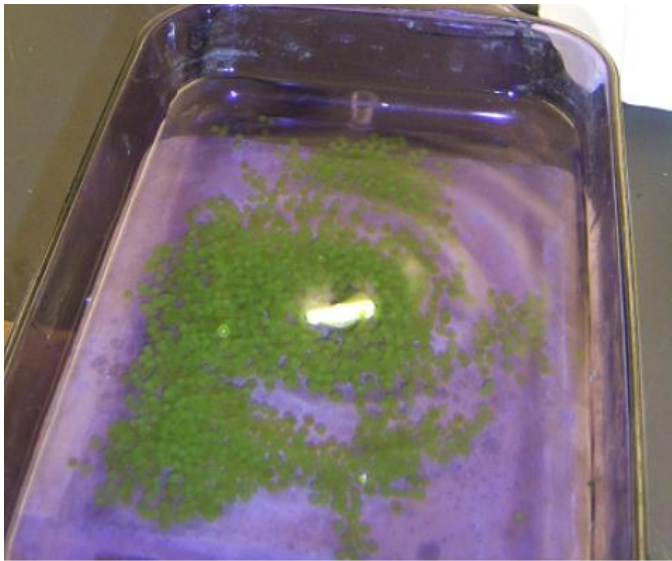


Figure 25. The experimental system designed for CO₂ consumption measurements of immobilized algae beads.

A- CO₂ tank

B- air tank

C- gas mixing chamber

D- CO₂ influent gas sample port*

E- Tygon tubes containing immobilized algae

F- CO₂ effluent gas sample port*

* - CO₂ measured by gas chromatography

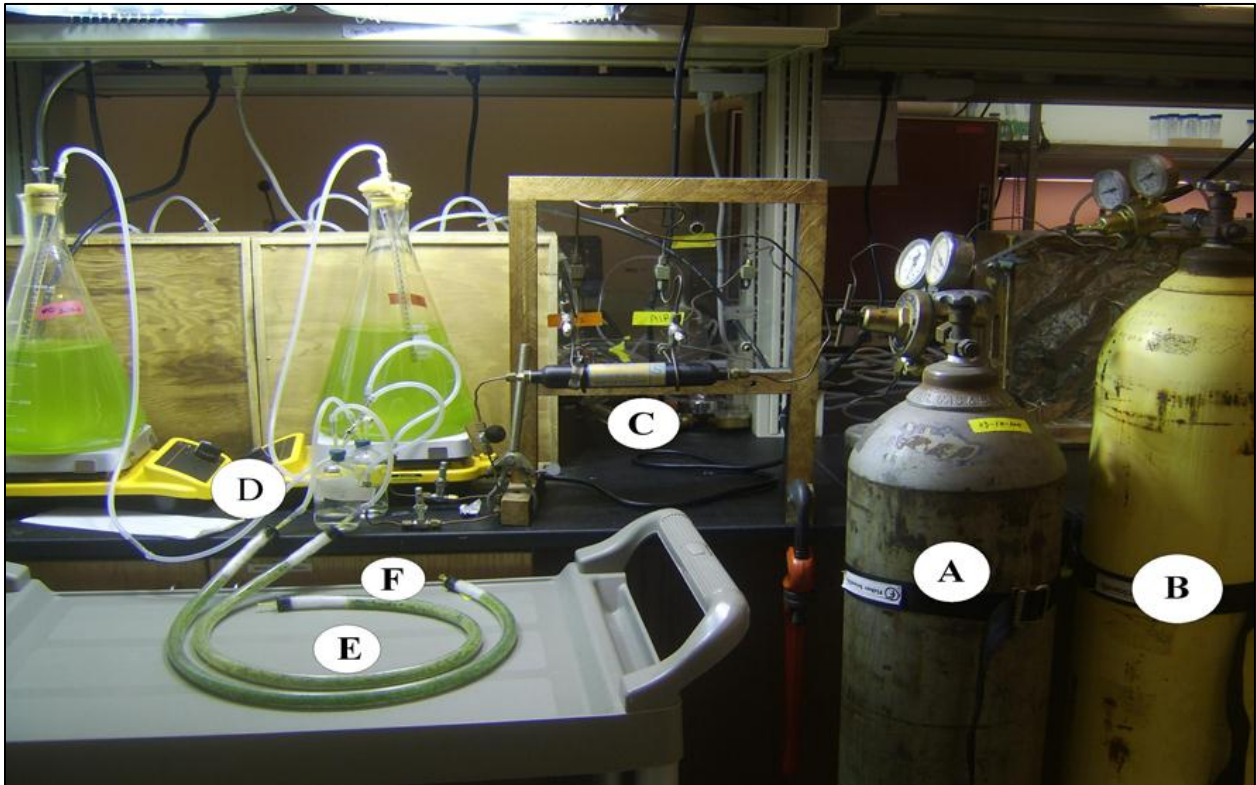


Figure 26. Consumption of 12% atmospheric CO₂ and growth over a 34-hrs period by *C. vulgaris* and *C. reinhardtii* using algal inoculum grown initially at ambient (0.038%) CO₂, 3% CO₂, 6% CO₂, 9% CO₂, and 12% CO₂.

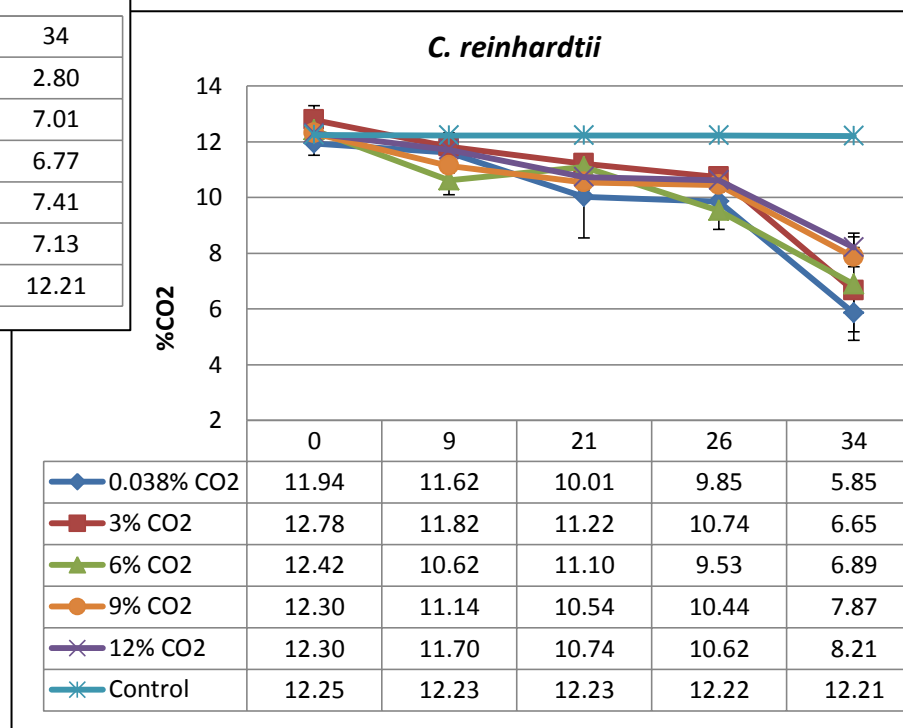
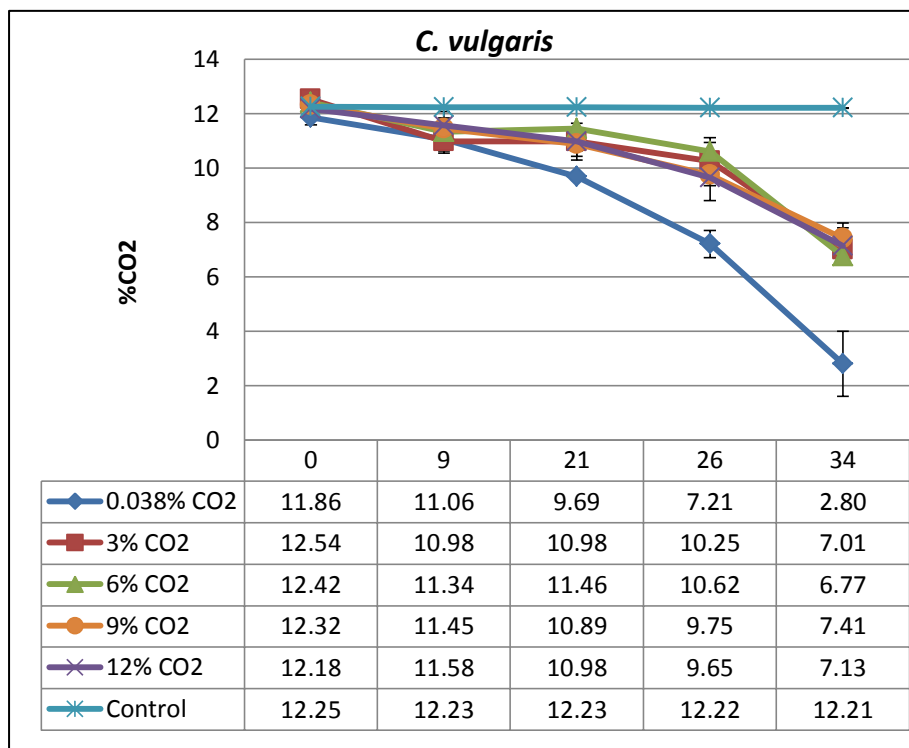


Figure 27. Growth of *C. vulgaris* and *C. reinhardtii* in serum bottle filled with 12% atmospheric CO₂ using algal inoculum grown initially at ambient (0.038%) CO₂, 3% CO₂, 6% CO₂, 9% CO₂, and 12% CO₂

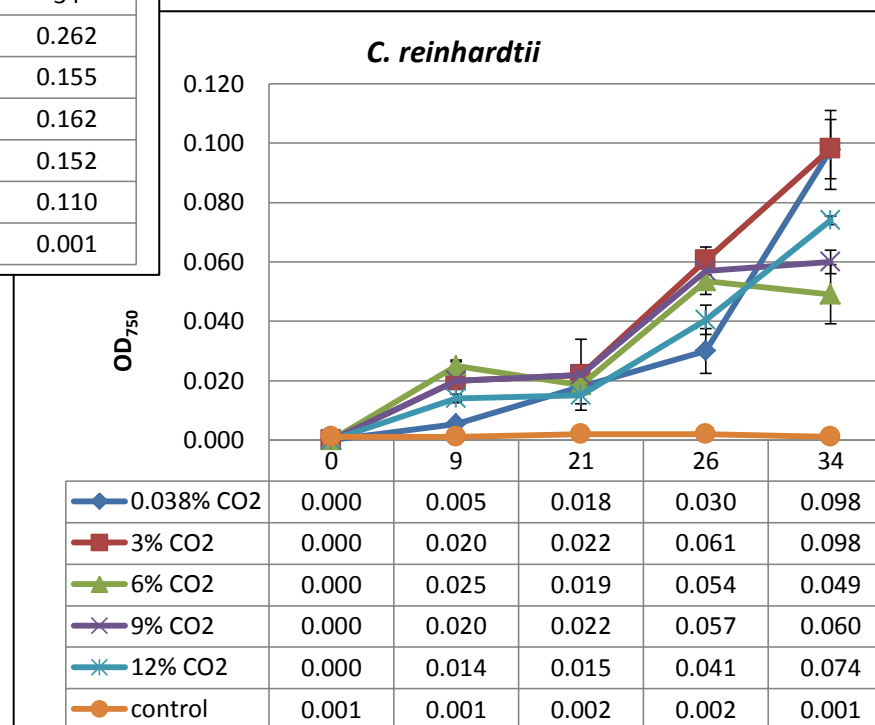
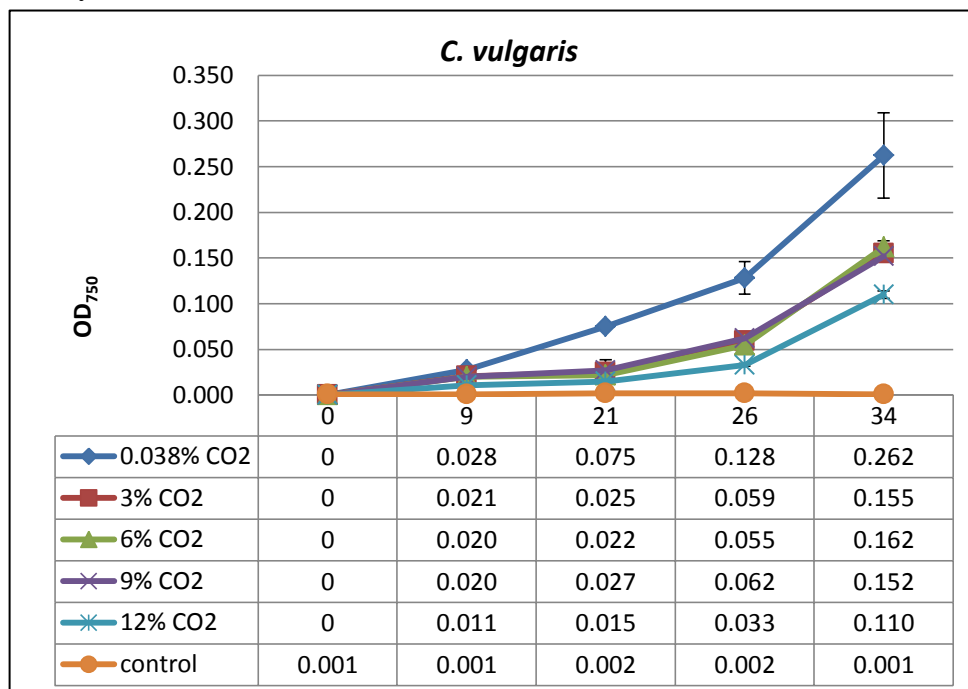


Table 24. CO₂ fixation rates (mg_{CO2} L⁻¹ day⁻¹) of 12%CO₂ atmospheric *C. vulgaris* and *C. reinhardtii* cells grown at ambient, 3%, 6%, 9%, and 12% CO₂.

Microalgal species	CO ₂ (%)	CO ₂ consumption rate (mg L ⁻¹ d ⁻¹)	Reference
<i>C. vulgaris</i>	0.038	278	This work
<i>C. vulgaris</i>	3	149	This work
<i>C. vulgaris</i>	6	152	This work
<i>C. vulgaris</i>	9	146	This work
<i>C. vulgaris</i>	12	152	This work
<i>C. reinhardtii</i>	0.038	176	This work
<i>C. reinhardtii</i>	3	169	This work
<i>C. reinhardtii</i>	6	154	This work
<i>C. reinhardtii</i>	9	121	This work
<i>C. reinhardtii</i>	12	117	This work
<i>Chlorella</i> sp.	10	1767	Sung <i>et al.</i> (1999)
<i>Chlorella</i> sp.	20	1316	Sakai <i>et al.</i> (1995)
<i>C. vulgaris</i>	0.038	75	Scragg <i>et al.</i> (2002)
<i>C. vulgaris</i>	10	612	Jin <i>et al.</i> (2006)
<i>C. vulgaris</i>	0.8–1	6240	Cheng <i>et al.</i> (2006)
<i>C. vulgaris</i>	0.09	3450	Fan <i>et al.</i> (2008)
<i>Chlorella</i> sp.	2	857	Chiu <i>et al.</i> (2008)
<i>Chlorella</i> sp.	10	717	Chiu <i>et al.</i> (2009)
<i>Chlorella</i> sp.	5	700	Ryu <i>et al.</i> (2009)
<i>C. vulgaris</i>	5	251	Sydney <i>et al.</i> (2010)

Table 25. Analysis of variance (ANOVA) table of CO₂ uptake in liquid culture after 34 hrs.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	9	31.1224	5.2915	0.0078
Algae type	1	2.5590	3.9020	0.0765
CO ₂ growth	4	23.4958	8.9567	0.0024
Algae type* CO ₂ growth	4	5.1775	1.9737	0.1747
Error	10	6.5582		
Total	19	37.7906		

Table 26. Analysis of variance (ANOVA) table of OD₇₅₀ in liquid culture after 34 hrs.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	9	0.0607	14.8536	0.0001
Algae type	1	0.0400	88.1924	<.0001
CO ₂ growth	4	0.0132	7.3160	0.0051
Algae type* CO ₂ growth	4	0.0074	4.0564	0.0330
Error	10	0.0045		
Total	19	0.0652		

Figure 28. Consumption of 3%, 6%, and 9% influent CO₂ flowing through immobilized *C. vulgaris* algal beads.

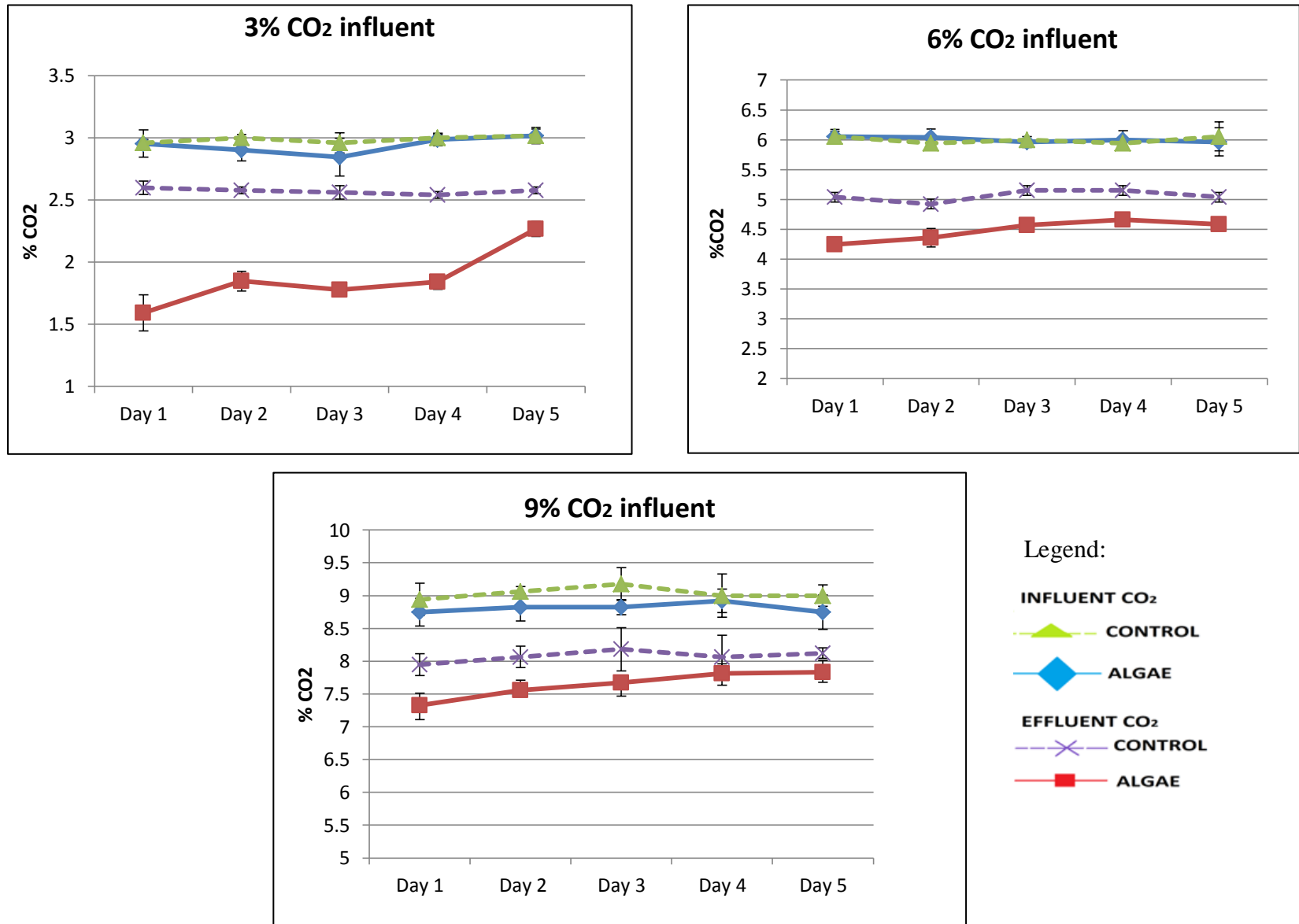


Figure 29. Consumption of 3%, 6%, and 9% influent CO₂ flowing through immobilized *C. reinhardtii* algal beads.

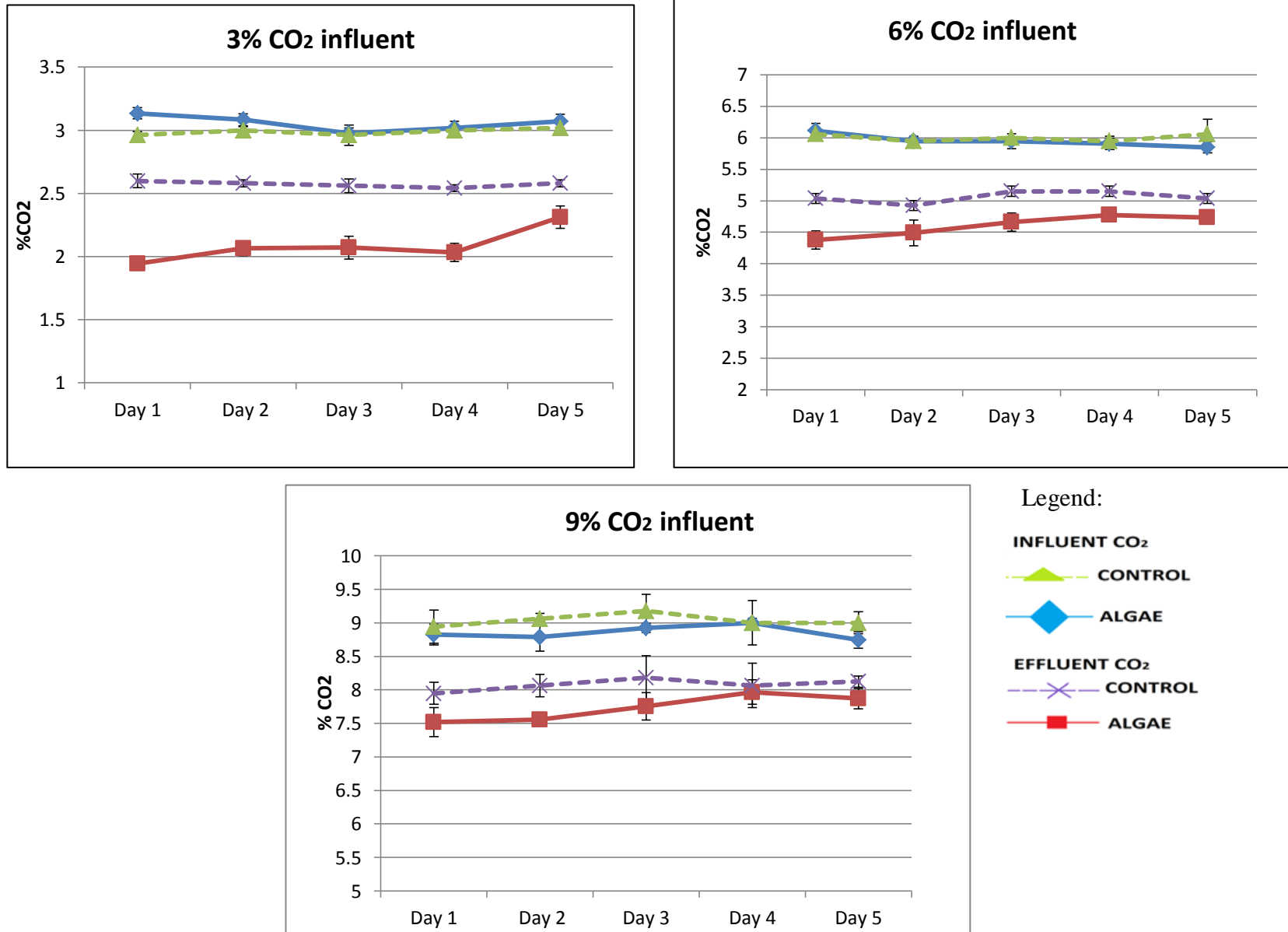


Figure 30. *C. vulgaris* % CO₂ reduction over a 5-day period at different %CO₂ influent.

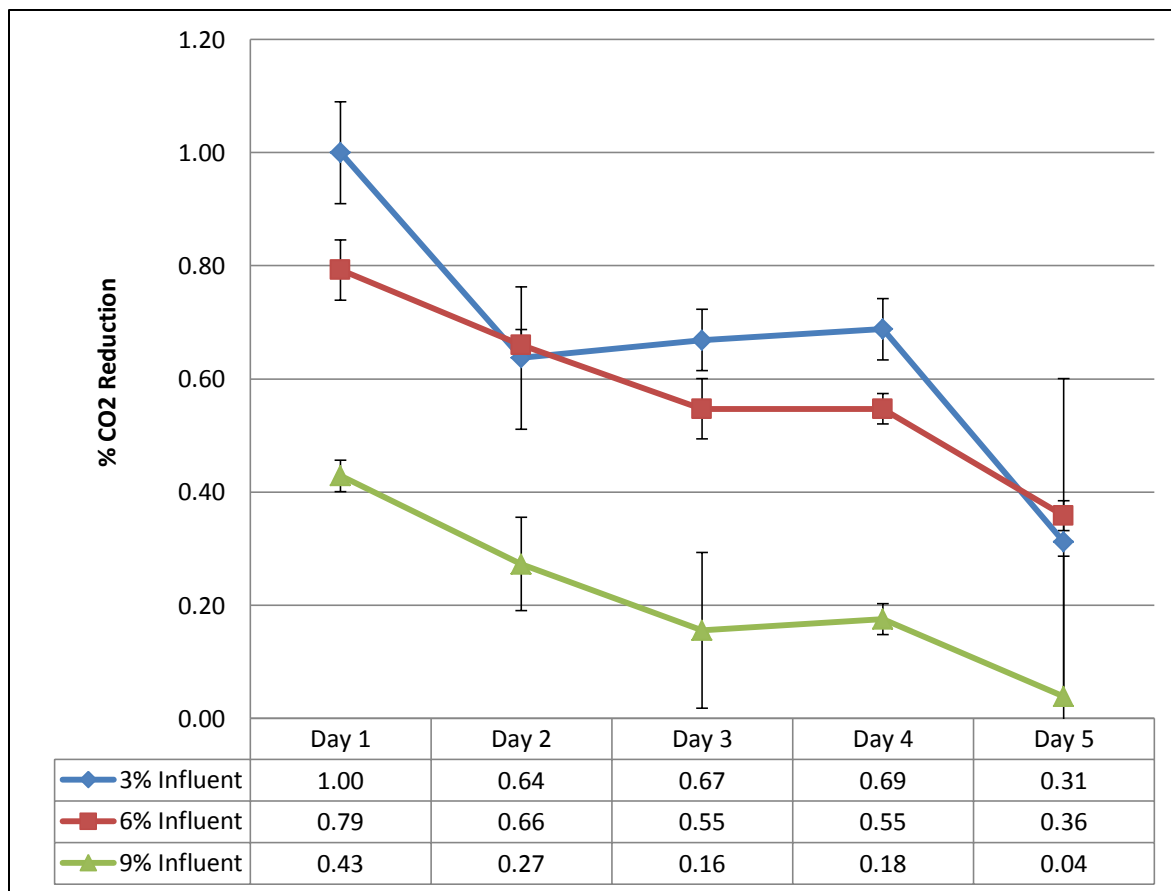


Figure 31. *C. reinhardtii* % CO₂ reduction over a 5-day period at different %CO₂ Influent.

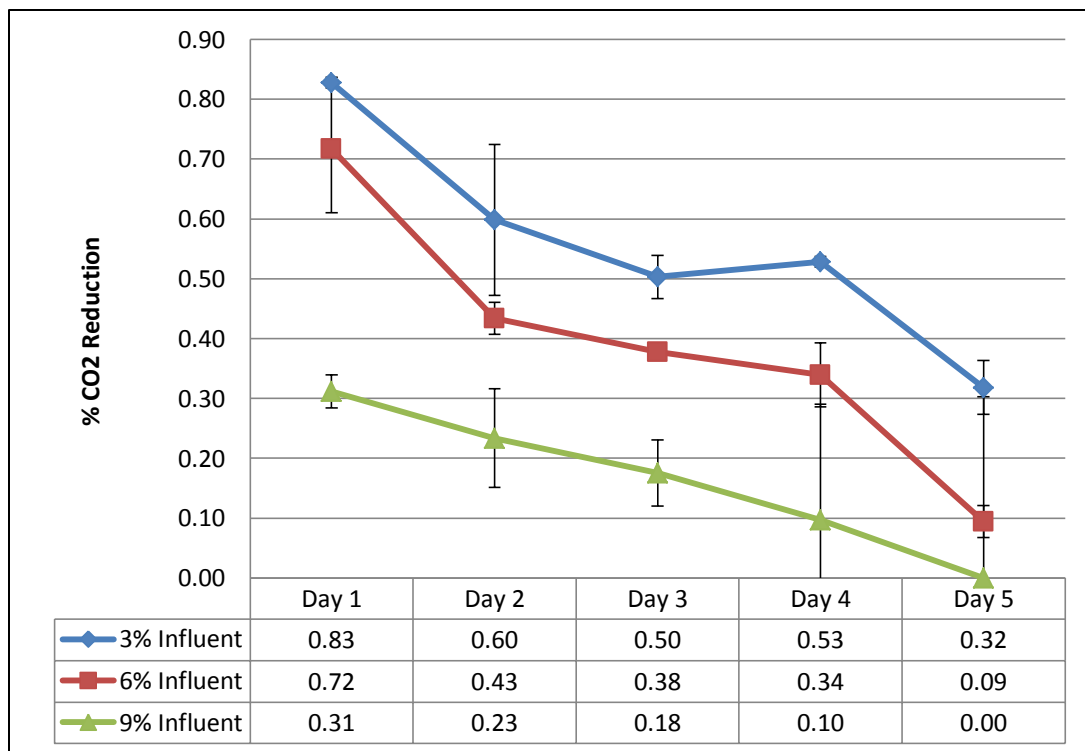


Table 27. Analysis of variance (ANOVA) table of %CO₂ reduction by algal beads on day one.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	5	0.6722	34.6232	0.0002
Algae type	1	0.0456	11.7511	0.0140
% CO ₂ Influent	2	0.6216	80.0365	<.0001
Algae type* % CO ₂ Influent	2	0.0050	0.6459	0.5571
Error	6	0.0233		
Total	11	0.6956		

Table 28. Analysis of variance (ANOVA) table of %CO₂ reduction by algal beads on day five.

Source of Variation	dF	Sum of Squares	F Ratio	Prob > F
Model	5	0.1613	1.4750	0.3222
Algae type	1	0.0225	1.0305	0.3492
% CO ₂ Influent	2	0.0910	2.0812	0.2058
Algae type* % CO ₂ Influent	2	0.04771	1.0911	0.3943
Error	6	0.1312		
Total	11	0.2925		

Table 29. Influent and effluent CO₂ concentration of control column and column containing algal beads.

A) Control column

%CO ₂	3% CO ₂		6% CO ₂		9% CO ₂	
Day	Influent	Effluent	Influent	Effluent	Influent	Effluent
Day 1	2.9618 ± 0.03	2.5987 ± 0.05	6.0566 ± 0.08	5.0377 ± 0.08	8.9416 ± 0.25	7.9481 ± 0.17
Day 2	3.0000 ± 0.03	2.5796 ± 0.08	5.9434 ± 0.08	4.9245 ± 0.08	9.0584 ± 0.08	8.0649 ± 0.17
Day 3	2.9618 ± 0.08	2.5605 ± 0.05	6.0000 ± 0.01	5.1509 ± 0.08	9.1753 ± 0.25	8.1818 ± 0.33
Day 4	3.0000 ± 0.03	2.5414 ± 0.08	5.9434 ± 0.08	5.1509 ± 0.08	9.0000 ± 0.33	8.0649 ± 0.33
Day 5	3.0191 ± 0.05	2.5796 ± 0.08	6.0566 ± 0.24	5.0377 ± 0.08	9.0000 ± 0.17	8.1234 ± 0.08

B) *C. vulgaris* column

%CO ₂	3% CO ₂		6% CO ₂		9% CO ₂	
Day	Influent	Effluent	Influent	Effluent	Influent	Effluent
Day 1	2.9554 ± 0.11	1.5923 ± 0.15	6.0566 ± 0.11	4.2453 ± 0.07	8.7468 ± 0.21	7.3247 ± 0.19
Day 2	2.9045 ± 0.09	1.8471 ± 0.08	6.0377 ± 0.14	4.3585 ± 0.15	8.8247 ± 0.21	7.5584 ± 0.16
Day 3	2.8471 ± 0.15	1.7771 ± 0.04	5.9622 ± 0.09	4.5660 ± 0.06	8.8247 ± 0.11	7.6753 ± 0.09
Day 4	2.9873 ± 0.05	1.8408 ± 0.06	6.0000 ± 0.15	4.6604 ± 0.09	8.9221 ± 0.18	7.8117 ± 0.15
Day 5	3.0191 ± 0.07	2.2675 ± 0.06	5.9623 ± 0.24	4.5849 ± 0.10	8.7468 ± 0.26	7.8312 ± 0.18

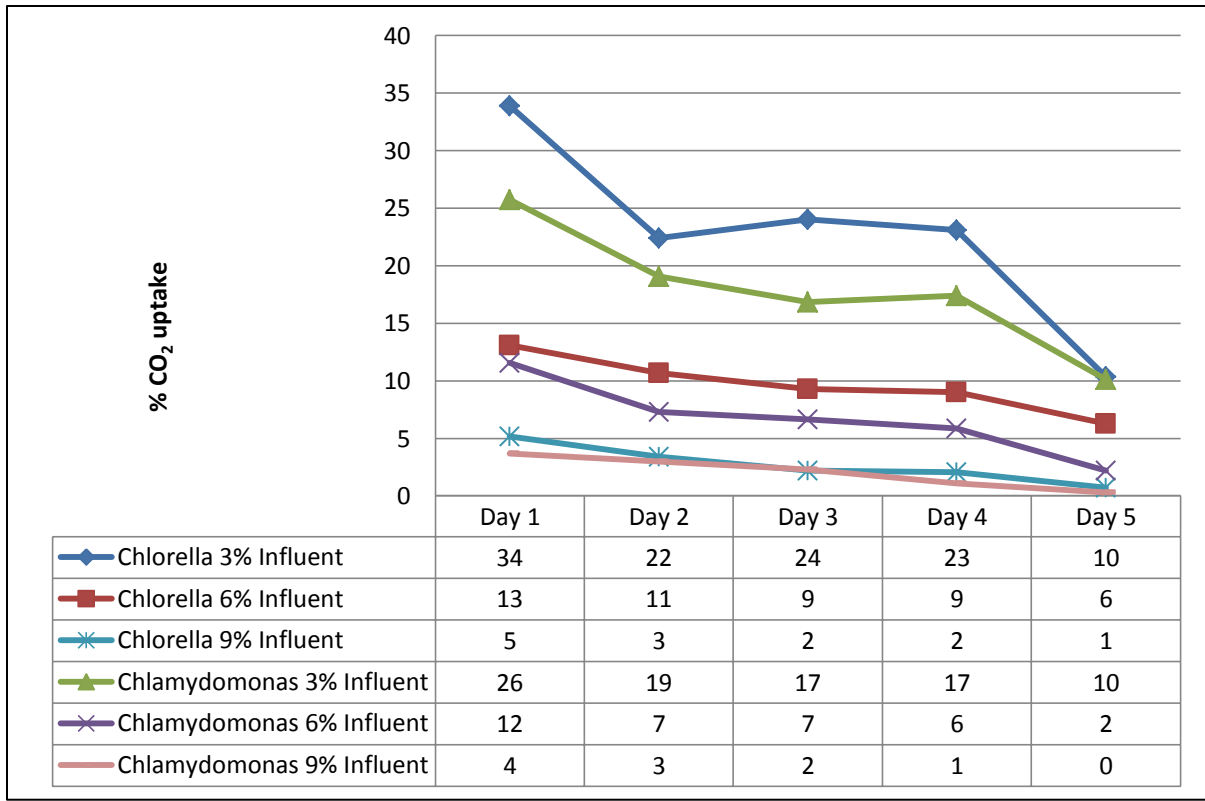
C) *C. reinhardtii* column

%CO ₂	3% CO ₂		6% CO ₂		9% CO ₂	
Day	Influent	Effluent	Influent	Effluent	Influent	Effluent
Day 1	3.1338 ± 0.04	1.9427 ± 0.04	6.0566 ± 0.11	4.2453 ± 0.07	8.8247 ± 0.16	7.5195 ± 0.21
Day 2	3.0828 ± 0.05	2.0637 ± 0.06	6.0377 ± 0.14	4.3585 ± 0.15	8.7857 ± 0.21	7.5584 ± 0.09
Day 3	2.9745 ± 0.04	2.0701 ± 0.09	5.9623 ± 0.09	4.5660 ± 0.07	8.9221 ± 0.07	7.7532 ± 0.21
Day 4	3.0191 ± 0.05	2.0318 ± 0.07	6.0000 ± 0.15	4.6604 ± 0.07	9.0000 ± 0.07	7.9675 ± 0.18
Day 5	3.0701 ± 0.06	2.3121 ± 0.09	5.9623 ± 0.24	4.5849 ± 0.10	8.7468 ± 0.12	7.8701 ± 0.15

Table 30. Rates of CO₂ uptake (mg L⁻¹ day⁻¹) by alginate-immobilized *C. vulgaris* and *C. reinhardtii* at 3%, 6% and 9% CO₂ influent.

%CO ₂ Influent	3% CO ₂		6% CO ₂		9% CO ₂	
	<i>C. vulgaris</i>	<i>C. reinhardtii</i>	<i>C. vulgaris</i>	<i>C. reinhardtii</i>	<i>C. vulgaris</i>	<i>C. reinhardtii</i>
Day 1	40.28	35.28	53.65	51.41	42.12	38.66
Day 2	31.32	30.18	49.74	43.03	37.50	36.35
Day 3	31.69	26.79	41.35	38.00	34.04	34.62
Day 4	33.96	29.24	39.68	33.53	32.89	30.578
Day 5	22.26	22.45	40.79	32.97	27.12	25.96

Figure 32. *C. vulgaris* and *C. reinhardtii* % CO₂ uptake at different influent CO₂%.



Appendix 1.

Fig. 1. *C. vulgaris* standard curve for conversion of OD₇₅₀ values to cell numbers.

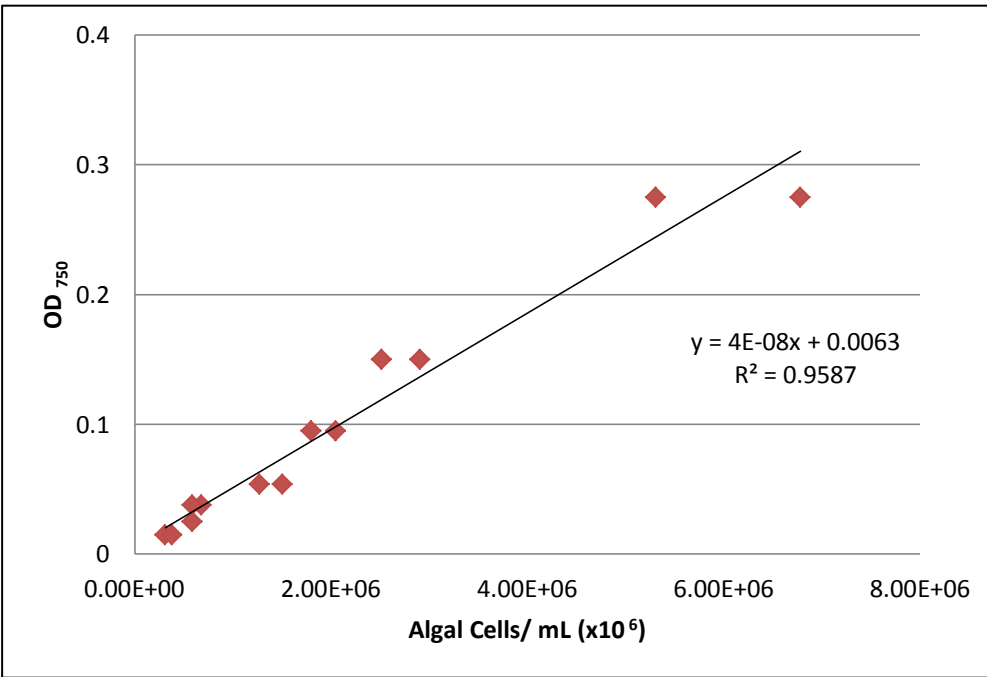


Fig. 2. *C. reinhardtii* standard curve for conversion of OD₇₅₀ values to cell numbers.

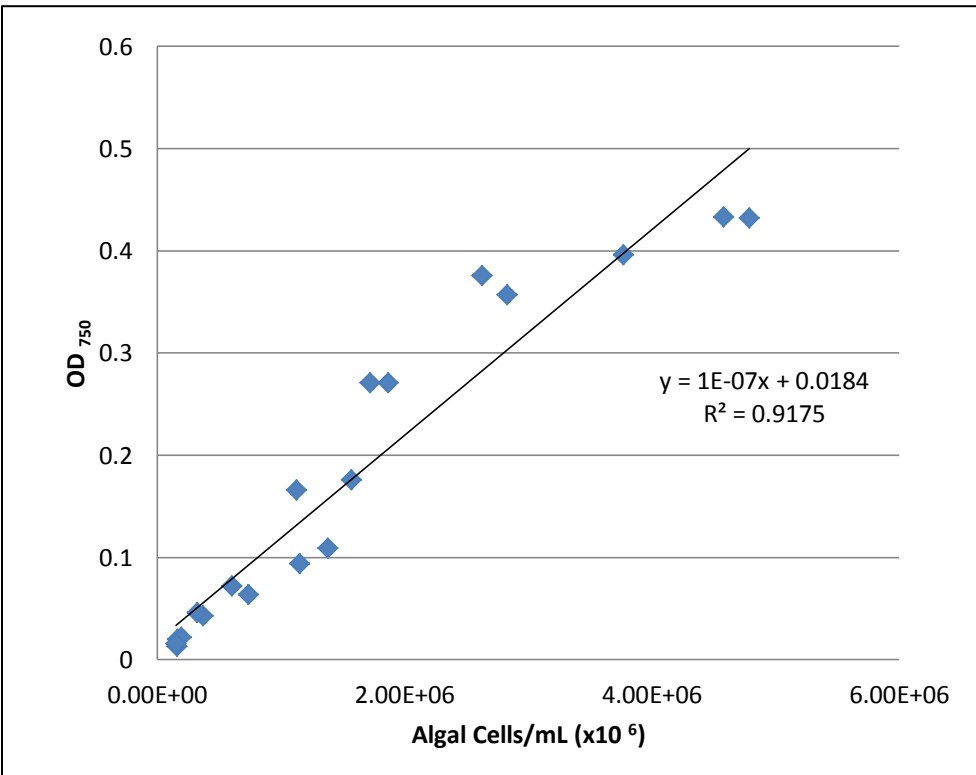


Table 1. Conversion of OD₇₅₀ values to cell numbers.

Experiments	Inoculum OD ₇₅₀	# Cell mL ⁻¹	
		<i>C. vulgaris</i>	<i>C. reinhardtii</i>
Chapter 1 (varying growth medium, pH, light, and CO ₂ concentration)	0.250	6.09E+06	2.32E+06
Chapter 2 Consumption of CO ₂ in liquid culture	0.050	1.09E+06	0.32E+06
Chapter 2 Consumption of CO ₂ flowing through immobilized algal beads	2.200	5.48E+07*	2.18E+07*

* # cells cm⁻³