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# Sulfidogenic and metal reducing activities of Desulfuromonas genus bacteria under the influence of copper chloride 

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The selection of strains isolated from technogenically altered ecotopes and resistant to contamination, capable of metabolizing a wide range of pollutants is a task highly relevant for creation of new methods for environmental purification. Sulphur-reducing bacteria of the Desulfuromonas genus carry out dissimilatory reduction not only of $S^{0}$ but also oxidized forms of metals. Intensity of anaerobic respiration of microorganisms in polluted environments is determined by level of their adaptation to stress factors, in particular, copper (II) compounds. The aim of this work was to investigate the influence of copper (II) chloride on $\mathrm{H}_{2} \mathrm{~S}$ production by Desulfuromonas sp. strains isolated by us from Yavorivske Lake, to determine the efficiency of $\mathrm{Cu}^{2+}$ precipitation by hydrogen sulfide, to analyse the possibility of usage by bacteria of $\mathrm{CuCl}_{2}$ as an electron acceptor of anaerobic respiration and to study the influence of $\mathrm{Cu}^{2+}$ on usage by these microorganisms of ferric (III) citrate, potassium dichromate or manganese (IV) oxide as electron acceptors. Bacteria were grown under anaerobic conditions in Kravtsov-Sorokin medium. To study the influence of $\mathrm{Cu}^{2+}$ on production by bacteria of $\mathrm{H}_{2} \mathrm{~S}$, their cells were incubated with $\mathrm{CuCl}_{2}(0.5-4.0 \mathrm{mM})$, washed and cultivated in a medium with $\mathrm{S}^{0}$. To determine the level of $\mathrm{Cu}^{2+}$ binding by $\mathrm{H}_{2} \mathrm{~S}$, produced by bacteria, cells were grown in a medium with $\mathrm{CuCl}_{2}(0.5-4.0 \mathrm{mM})$ and $\mathrm{S}^{0}$. To investigate the ability of bacteria to use copper (II) ions as electron acceptors, they were cultivated in a medium with $\mathrm{CuCl}_{2}$ ( $1.74-10.41 \mathrm{mM}$ ). To study the influence of $\mathrm{Cu}^{2+}$ on usage by bacteria of metal compounds as electron acceptors, their cells were incubated with $\mathrm{CuCl}_{2}(0.5-4.0 \mathrm{mM})$, washed and cultivated in media with $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}(1.74-10.41 \mathrm{mM})$. Biomass was determined by the turbidimetric method. In the cultural liquid the content of $\mathrm{H}_{2} \mathrm{~S}$ was determined quantitatively by the spectrophotometric method, qualitatively - presence of $\mathrm{Cu}^{2+}$. Content of CuS in the growth medium was determined by weight method. Desulfuromonas sp. bacteria was revealed to be resistant to $2.0-2.5 \mathrm{mM}$ copper (II) ions. Under the influence of $3.0-$ $4.0 \mathrm{mM} \mathrm{CuCl}_{2}$ in the incubation mixture, sulfidogenic activity of bacteria decreased more than twice. The efficiency of $\mathrm{Cu}^{2+}$ binding in form of CuS by $\mathrm{H}_{2} \mathrm{~S}$ produced by bacteria reached $97.3-100.0 \%$ at presence in the medium with $\mathrm{S}^{0}$ of up to $1.5 \mathrm{mM} \mathrm{CuCl}_{2}$. Bacteria used $\mathrm{CuCl}_{2}(1.74-10.41 \mathrm{mM})$ as an electron acceptor in the process of anaerobic respiration. The addition of $2.5-3.0 \mathrm{mM}$ $\mathrm{CuCl}_{2}$ to the incubation mixture caused inhibition of metal reducing activity of cells, growth of all strains in media with 1.74 10.41 mM ferric (III) citrate, potassium dichromate or manganese (IV) oxide as electron acceptors decreased by 2.6 times. Almost complete precipitation up to 1.5 mM copper (II) ions in form of CuS by $\mathrm{H}_{2} \mathrm{~S}$ produced by bacteria and ability to reduce up to $10.41 \mathrm{mM} \mathrm{CuCl}_{2}, \mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ in the process of anaerobic respiration indicates a high adaptation of the bacteria strains investigated by us to stress factors, in particular, the influence of $\mathrm{CuCl}_{2}$. We have proved the possibility of using Desulfuromonas sp. in biotechnologies for purification of environments with complex contamination from copper (II) compounds.

Keywords: sulphur-reducing bacteria; hydrogen sulfide; copper; heavy metals.

## Introduction

In nature, $\mathrm{Cu}\left(6.2 \times 10^{-2}\right.$ mass $\left.\%\right)$ occurs both in a free state and in the form of compounds. Much copper is concentrated in the earth's crust in the form of sulfides $\left(\mathrm{CuFeS}_{2}\right.$ - chalcopyrite; $\mathrm{Cu}_{2} \mathrm{~S}$ - copper sulfide, chalcocite; CuS - covellite). Copper is part of oxides $\left(\mathrm{Cu}_{2} \mathrm{O}-\right.$ cuprite) and carbonates $\left((\mathrm{CuOH})_{2} \mathrm{CO}_{3}\right.$ - malachite) (Betehtin, 2007). The total content of copper in the upper layers of soil varies within 5$50 \mathrm{mg} / \mathrm{kg}$ (Kucherjavyj, 2001). Excessive amounts of copper come into the environment from pesticides, in particular fungicides, sewage of galvanic plants. In the flooded quarries of sulphur deposits, waste water from industrial enterprises, waste heaps of coal mines, landfills of household and industrial waste, the content of Cu can grow by a thousand times (Frank \& Lushnikov, 2006; Diakiv et al., 2016). In the technogenic reservoir which was formed in the place of the Yavoriv sulphur deposit quarry in recent years the content of copper (II) ions at depths of $0-70 \mathrm{~m}$ does not exceed $0.017-0.021 \mathrm{mg} / \mathrm{l}$ (Moroz et al., 2017).

The toxic effect of copper on cells of living organisms is due to its interaction with the SH-groups of amino acids of structural proteins and enzymes, which leads to their denaturation or inactivation (Ladomersky \&

Petris, 2015; Martínez-Bussenius et al., 2017). In the cases of chronic intoxication of humans with copper there can be frustrations of the nervous system, disorders of the kidneys and liver functions, dermatitis. Oxidized compounds of $\mathrm{Cu}, \mathrm{Hg}$ and Cr are considered to be the most toxic to microorganisms, compared to other heavy metals (Tashirev et al., 2007; Cidre et al., 2017). Heavy metals pro-oxidants, such as $\mathrm{Cu}^{2+}$, which are characterized by high oxidation-reduction potential $\left(E_{0}{ }^{\prime}\right)$, are toxic and inhibit the growth of microorganisms (Tashirev et al., 2008; Richter et al., 2012; Viti et al., 2014). Due to their presence in the cell, the microbial metabolism changes in direction of metal reduction instead of electron acceptors with lower $E_{0}{ }^{\prime}\left(\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}, \mathrm{SO}_{4}^{2-}, \mathrm{S}_{2} \mathrm{O}_{3}{ }^{2-}, \mathrm{SO}_{3}{ }^{2-}\right.$, $\mathrm{S}^{0}, \mathrm{HCO}_{3}^{-}$) reduction, which leads to generation of chemically active intermediates, oxygen radicals, energy exhaustion of cells and their death (Gescher \& Kappler, 2012; Ladomersky \& Petris, 2015; Simonte et al., 2017). Copper as a trace element at low concentrations is required by microorganisms for oxidation-reduction processes, stabilization of molecules in the case of electrostatic interactions, regulation of osmotic pressure, as a cofactor of metal proteins and some enzymes (Kumar et al., 2010). Copper influences the membrane $\mathrm{Na}^{+} / \mathrm{K}^{+}$ATP-ases, the structure and functions of nucleic acids, and the synthesis of phospholipids. It is
present in the composition of enzymes and enzyme complexes $(\mathrm{Cu}, \mathrm{Zn}-$ superoxide dismutase, lysiloxidase, doramine- $\beta$-hydroxylase, ascorbate oxidase, cytochrome $c$-oxidase, glutamyl transferase, hyponitrite reductase and nitrogen oxide reductase) (Lengeler et al., 2005). At high concentrations, copper represses $\alpha$-ketoglutarate dehydrogenase, inhibits nitrate reductase, vitamin $\mathrm{B}_{12}$ synthesis, processes of photosynthesis, fermentation and respiration in microorganisms due to reduced content of cytochromes $b$ and $c$ (Segin et al., 2016; Martínez-Bussenius et al., 2017). Cop transport system of copper (chromosomal cop-operon) and its regulation in Enterococcus hirae has been described (Solioz \& Stoyanov, 2003; Winkelmann, 2008). Copper resistance, which is determined by plasmid genes, is described in Xanthomonas (Behlau et al., 2017; Richard et al., 2017a; Gochez et al., 2018), Cupriavidus (Gillan et al., 2017), Pseudomonas, Escherichia and other genera of bacteria (Silver \& Walderhaug, 1995; Richard et al., 2017b).

The selection of strains isolated from technogenically altered ecotopes and adapted to contamination, which are capable of metabolizing a wide spectrum of pollutants is a particularly relevant task in terms of establishing the mechanisms of their resistance (Limcharoensuk et al., 2015; Si et al., 2015; Dey et al., 2016) and creation of new methods for environmental purification (Iwahori et al., 2014; Mustapha \& Halimoon, 2015). Sulfidogenic sulphur-reducing bacteria which reduce elemental or polysulfide forms of sulphur by action of sulphur reductase or polysulfide reductase, localized in cytoplasmic membrane and bound with hydrogenase by cytochromes or quinones (Hedderich et al., 1999), with the formation of hydrogen sulfide, attract attention of biotechnologists as potential agents of purification of environments contaminated by metal compounds, because as a result of interaction of $\mathrm{H}_{2} \mathrm{~S}$ with bivalent metal ions their insoluble sulfides are formed and thus they are removed from the natural cycle of elements (Diakiv et al., 2017; Kiran et al., 2017). Previously we have shown that bacteria strains of Desulfuromonas genus, isolated from Yavorivske Lake, have high metal reducing activity and may except sulphur using $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ as electron acceptors in the process of anaerobic destruction of organic compounds (Moroz et al., 2014; Moroz et al., 2017). Metal reducing bacteria with involvement of metal reductases (multiheme $c$-type cytochromes) (Kozlova et al., 2008; Richter et al., 2012) enzymatically reduce Fe (III), Cr (VI), Mn (IV), U (VI), Tc (VII), Pd (II), V (V), Mo (VI), Cu (II) etc., which is one of the mechanisms of protection of cells from their toxic effects (Gescher \& Kappler, 2012; Viti et al., 2014; Maslovska \& Hnatush, 2015). The components of the Mtr (metal reducing) electron transfer respiratory chain containing the complex of c-type cytochromes are best studied today in bacteria of the Shewanella and Geobacter genera (Gescher \& Kappler, 2012; Breuer et al., 2015; Simonte et al., 2017). Membrane-bound metal reductases (tetra- and decaheme c-type cytochromes) in S. frigidimarina are localized in the periplasm through which electrons from the cytoplasm from the carbon compounds metabolism reactions through the inner and outer membranes and periplasm are transmitted outside the cells, where in fact the soluble or insoluble metal compounds are reduced (Lengeler et al., 2005; Kozlova et al., 2008; Richter et al., 2012). The reduction of cations or oxygen-containing metal anions by metal reductases is carried out outside the cell, which is accompanied by release of a significant amount of electrons into the medium, therefore bacteria with exoelectrogenic properties are considered as possible anode biocatalysts in microbial fuel cells (Bilyy et al., 2014; Simonte et al., 2017).

Studies of the influence of $\mathrm{Cu}^{2+}$ on $\mathrm{H}_{2} \mathrm{~S}$ biogenesis by bacteria of Desulfuromonas sp. are important for development of effective and profitable biological methods of regulating their level in transformed biotopes. Investigations of the ability of sulphur-reducing bacteria to reduce $\mathrm{Cu}^{2+}$ are necessary for deepening understanding of mechanisms of their adaptation to existence in environments contaminated by copper (II) compounds. Studies of the influence of $\mathrm{CuCl}_{2}$ on the ability of bacteria to reduce other metals will allow us to evaluate their resistance to copper (II) cations and the possibility of their application in environmental remediation technologies. Therefore, the purpose of this work was to investigate the influence of copper (II) chloride on $\mathrm{H}_{2} \mathrm{~S}$ production by sulphur-reducing bacteria of the Desulfuromonas genus isolated from Yavorivske Lake, to determine the efficiency of copper (II) cations precipitation by hydrogen sulfide, to analyse the possibility of usage by
bacteria of $\mathrm{CuCl}_{2}$ as an electron acceptor of anaerobic respiration and to study the influence of $\mathrm{Cu}^{2+}$ on usage by bacteria of ferric (III) citrate, potassium dichromate or manganese (IV) oxide as electron acceptors to assess the biotechnological potential of the investigated strains and to expand theoretical knowledge about changes in metabolism of sulphurreducing bacteria under the influence of stress factors.

## Materials and methods

Sulphur-reducing bacteria Desulfuromonas acetoxidans IMV B-7384, isolated by us earlier from Yavorivske Lake, are identified and stored at the Depository of D. K. Zabolotny Institute of Microbiology and Virology of the NAS of Ukraine and in the collection of Microbiology Department of Ivan Franko National University of L'viv (Gudz et al., 2013). Bacteria Desulfuromonas sp. Yavor-5 and Desulfuromonas sp. Yavor-7, isolated by us from the same lake, are identified and stored in the collection of Microbiology Department of Ivan Franko National University of L’viv (Moroz et al., 2013).

The bacteria were grown in Kravtsov-Sorokin medium (Gudz et al., 2014) without Mohr's salt and without sulfate ions with sulphur of the following composition (g/l): $\mathrm{NaH}_{2} \mathrm{PO}_{4} \times 12 \mathrm{H}_{2} \mathrm{O}-0.84, \mathrm{~K}_{2} \mathrm{HPO}_{4}-0.5$, $\mathrm{NH}_{4} \mathrm{Cl}-0.16, \mathrm{MgCl}_{2} \times 6 \mathrm{H}_{2} \mathrm{O}-0.10, \mathrm{NaC}_{3} \mathrm{H}_{5} \mathrm{O}_{3}-2.0$. Before bacteria seeding, 0.05 ml of $\mathrm{Na}_{2} \mathrm{~S} \times 9 \mathrm{H}_{2} \mathrm{O}(1 \%)$ sterile solution was added to the medium. A sterile 10 N NaOH solution was used to provide pH of the medium to 7.2 . Cells were added to the medium in quantity of $10 \mathrm{vol} . \%$ to initial concentration $10^{8} \mathrm{CFU} / \mathrm{ml}(0.05 \mathrm{~g} / \mathrm{l})$. The sulphur was sterilized separately $(0.5 \mathrm{~atm})$ and placed in the medium at concentration of not less than $0.11 \mathrm{~g} / \mathrm{l}\left(3.47 \mathrm{mM}\right.$ - concentration of $\mathrm{SO}_{4}{ }^{2-}$ in medium of standard composition). Bacteria were grown for 10 days in tubes of 25 ml volume under anaerobic conditions at temperature of $30^{\circ} \mathrm{C}$.

Biomass was determined by the turbidimetric method using KFK-3 at 340 nm wavelength and 3 mm optical path and calculated in $\mathrm{g} / \mathrm{l}$ using the formula: $\mathrm{C}=\left(\mathrm{E}_{340} \cdot \mathrm{n}\right) / \mathrm{K}$, where E - extinction at 340 nm wavelength, n - dilution, times, K - coefficient of recalculation, obtained by calibration curve of dependence of extinction from the mass of dry cells, determined by weight method and equal to 0.72 (Gudz et al., 2014). The concentration of hydrogen sulfide in the cultural liquid, separated from cells by centrifugation ( $4025 \mathrm{~g}, 20 \mathrm{~min}$ ), was determined spectrophotometrically by the formation of methylene blue due to interaction of N,N-dimethyl-n-phenylenediamine dihydrochloride and hydrogen sulfide (Gudz et al., 2014).

To determine the influence of $\mathrm{CuCl}_{2}$ on growth and hydrogen sulfide production by bacteria cells, previously grown in medium with fumarate $\left(\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}\right)$ as electron acceptor $(3.47 \mathrm{mM})$ and cysteine $\left(\mathrm{C}_{3} \mathrm{H}_{7} \mathrm{NO}_{2} \mathrm{~S}\right)$ at concentration of $0.2 \mathrm{~g} / \mathrm{l}$ (Lengeler et al., 2005) to meet assimilation needs of bacteria in sulphur, were precipitated by centrifugation $(4025 \mathrm{~g}$, $20 \mathrm{~min})$, resuspended in $\mathrm{NaCl}(0.9 \%)$ sterile solution and incubated for one hour with sterile solution of $\mathrm{CuCl}_{2}$ at concentrations: 0 (control), $0.5,1.0,1.5,2.0,2.5,3.0,4.0 \mathrm{mM}$, under sterile conditions. Cells were precipitated by centrifugation, washed twice with $\mathrm{NaCl}(0.9 \%)$ solution and sown in medium with $\mathrm{S}^{0}$. After 10 days of growth the biomass and the content of hydrogen sulfide in the cultural liquid were determined.

The level of $\mathrm{Cu}^{2+}$ binding by $\mathrm{H}_{2} \mathrm{~S}$, produced by bacteria, was determined. For this, bacteria were grown in a medium with $\mathrm{S}^{0}$ and $\mathrm{CuCl}_{2}$ at concentrations: 0 (control), $0.5,1.0,1.5,2.0,2.5,3.0,4.0 \mathrm{mM}$. After 10 days of growth biomass was determined. The mixture of cells and CuS was precipitated by centrifugation $(4025 \mathrm{~g}, 20 \mathrm{~min})$, in the cultural liquid the presence of copper (II) cations was determined qualitatively (Harris, 2003) and hydrogen sulfide concentration quantitatively, the total concentration of which was calculated as sum of concentrations of free $\mathrm{H}_{2} \mathrm{~S}$ and bound in form of metal sulfide. The content of CuS was determined by weight method. For this, a mixture of cells and CuS was weighed; the mass of metal sulfide was calculated as difference between mass of mixture and dry cells (grown after incubation with $\mathrm{CuCl}_{2}$ ) and medium components. The relative concentration (\%) of bound by $\mathrm{H}_{2} \mathrm{~S}$ metal cations was calculated based on the ratio of molar concentrations of formed copper (II) sulfide and copper (II) ions, added into the medium at the beginning of bacteria cultivation, taking $\mathrm{CuCl}_{2}$ concentration as $100 \%$. To study the biomass accumulation by bacteria during usage of $\mathrm{CuCl}_{2}$ or $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}$ (control) as electron acceptors, cells
were sown in medium without $\mathrm{S}^{0}$ with $\mathrm{C}_{3} \mathrm{H}_{7} \mathrm{NO}_{2} \mathrm{~S}(0.2 \mathrm{~g} / \mathrm{l})$ as sulphur source. Sterile 1 M solutions of fumarate and copper (II) chloride were added into medium to their final concentrations that were $0.5,1.0,1.5,2.0$ and 3.0 times different from the standard electron acceptor content in Kravtsov-Sorokin medium ( $1.74,3.47,5.21,6.94,10.41 \mathrm{mM}$ ). Previously the bacteria were grown in a medium without $S^{0}$ with fumarate ( 3.47 mM ) and cysteine $(0.2 \mathrm{~g} / \mathrm{l})$ to the middle of the exponential growth phase. At 2, 4, 6, 8, 10 days biomass was determined. In the cultural liquid the presence of copper (II) cations was determined qualitatively (Harris, 2003).

To investigate the influence of $\mathrm{CuCl}_{2}$ on use by bacteria of $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}$, $\mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ as electron acceptors the cells, previously grown in a medium with $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}(3.47 \mathrm{mM})$ and $\mathrm{C}_{3} \mathrm{H}_{7} \mathrm{NO}_{2} \mathrm{~S}(0.2 \mathrm{gl})$ to the middle of exponential growth phase, were precipitated by centrifugation $(4025 \mathrm{~g}$, $20 \mathrm{~min})$, resuspended in sterile $\mathrm{NaCl}(0.9 \%)$ solution, under sterile conditions incubated for one hour with sterile $\mathrm{CuCl}_{2}$ solution at concentrations: 0 (control), $0.5,1.0,1.5,2.0,2.5,3.0,4.0 \mathrm{mM}$, precipitated by centrifugation, washed twice with $\mathrm{NaCl}(0.9 \%)$ solution and sown in a


Fig. 1. The influence of $\mathrm{CuCl}_{2}$ on biomass accumulation $(a, b, c)$ and hydrogen sulfide production $(d, e, f)$ by $D$. acetoxidans IMV B-7384 $(a, d)$, Desulfuromonas sp. Yavor-5 $(b, e)$ and Desulfuromonas sp. Yavor-7 $(c, f)$ after 10 days of growth in medium with $\mathrm{S}^{0} ; *-\mathrm{P}<0.05$

To study the level of copper (II) ions binding by produced by bacteria, hydrogen sulfide cells were grown in medium with $\mathrm{S}^{0}$ and $\mathrm{CuCl}_{2}$ at different concentrations. Strains of bacteria of Desulfuromonas genus proved to be resistant to $\mathrm{CuCl}_{2}$ influence at concentrations of $0.5-$ 2.5 mM both in the incubation mixture and in the cultivation medium. Under the influence of $3.0-4.0 \mathrm{mM} \mathrm{CuCl}_{2}$, the accumulation of biomass and the production of $\mathrm{H}_{2} \mathrm{~S}$ by sulphur-reducing bacteria in the medium with elemental sulphur decreased by 2.1-3.0 and 1.7-2.2 times, respectively (Fig. 1).

The cells of bacteria, incubated with $0.5-1.5 \mathrm{mM} \mathrm{CuCl}_{2}$, in 10 days produced not less than $1.8-2.0 \mathrm{mM} \mathrm{H}_{2} \mathrm{~S}$. Therefore, the efficiency of copper (II) ions binding in form of CuS by hydrogen sulfide produced by bacteria reached $97.3-100.0 \%$ at the presence in medium with $\mathrm{S}^{0}$ of $0.5-1.5 \mathrm{mM} \mathrm{CuCl}_{2}$ (Table 1). After addition during seeding into the medium of up to 1.5 mM copper (II) cations for 10 days they were completely binded with the $\mathrm{H}_{2} \mathrm{~S}$, produced by bacteria, which confir-
med the negative results of qualitative reactions on their presence in the cultural liquid. The binding level of copper (II) ions, added to medium at the beginning of cultivation at concentration of 2.0 mM , with the $\mathrm{H}_{2} \mathrm{~S}$ produced by bacteria was 1.3 times lower than at lower concentrations of $\mathrm{CuCl}_{2}$ in the medium and did not exceed $81.0 \%$, since it amount was not sufficient for full interaction with metal ions. After addition of more than $2.0 \mathrm{mM} \mathrm{Cu}{ }^{2+}$ into the medium, positive results of qualitative reactions indicated the presence of metal ions in the cultural liquid.

Consequently, it was found that under the influence of $3.0-4.0 \mathrm{mM}$ $\mathrm{CuCl}_{2}$, sulfidogenic activity of Desulfuromonas sp. decreased more than twice. The efficiency of copper (II) ions precipitation in form of CuS by $\mathrm{H}_{2} \mathrm{~S}$ produced by bacteria reached $97.3-100.0 \%$ at presence in the medium with $\mathrm{S}^{0}$ up to $1.5 \mathrm{mM} \mathrm{CuCl}_{2}$. The influence of $\mathrm{CuCl}_{2}$ on biomass accumulation by sulphur-reducing bacteria was studied. Cells were grown in media with cysteine as a source of sulphur and $\mathrm{CuCl}_{2}$ or $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}$ (control) at concentrations of $1.74,3.47,5.21,6.94,10.41 \mathrm{mM}$ (Fig. 2).


Fig. 2. Biomass accumulation by D. acetoxidans IMV B-7384 ( $a, d$ ), Desulfuromonas sp. Yavor-5 ( $b, e$ ) and Desulfuromonas sp. Yavor-7 $(c, f)$ during growth in medium with $\mathrm{CuCl}_{2}(a, b, c)$ or $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}(d, e, f)$ at different concentrations

With increasing of $\mathrm{CuCl}_{2}$ concentrations in the cultivation medium, the biomass of Desulfuromonas sp. strains decreased. The largest biomass of bacteria was accumulated for $8-10$ days in the medium with the lowest concentration of copper (II) ions -1.74 mM (up to $1.44 \mathrm{~g} / \mathrm{l}$ ), and the smallest - with the highest -10.41 mM (up to $1.22 \mathrm{~g} / \mathrm{l}$ ), which can be explained by the toxic effect of metal ions on bacteria cells. At presence in the medium of $\mathrm{CuCl}_{2}$ in all concentrations for 10 days, biomass of bacteria was 1.5-2.0 times lower than in the medium with fumarate. In contrast to the bacteria growth in the medium with $\mathrm{CuCl}_{2}$, with increasing concentrations of fumarate in the cultivation medium from 1.74 to 3.47 mM for 4-10 days a rise in biomass accumulation by bacteria of all strains was observed from 2.29 to $2.48 \mathrm{~g} / \mathrm{l}$. However, the highest biomass (up to $2.60 \mathrm{~g} / \mathrm{l}$ ) bacteria accumulated in the medium with fumarate at concentration of 5.21 mM . With further growth of fumarate concentrations in the medium from 6.94 to 10.41 mM , the biomass decreased and did not exceed $2.39 \mathrm{~g} / \mathrm{l}$, possibly due to limiting of bacteria growth by other factors in the medium. Copper (II) ions were found in the medium throughout all time of bacteria cultivation (Table 2), which indicates its incomplete use, apparently, due to high toxicity.



Thus, it has been found that bacteria Desulfuromonas sp. used copper (II) ions as electron acceptors in the process of anaerobic respiration at concentrations of $1.74-10.41 \mathrm{mM} \mathrm{CuCl}_{2}$ in the medium. A $1.5-2.0$ times lower biomass yield was found during use by bacteria of copper (II) chloride, compared with use of fumarate in all investigated concentrations of electron acceptors in the medium. The highest biomass (up to $1.44 \mathrm{~g} / \mathrm{l})$ bacteria accumulated in the medium with $1.74 \mathrm{mM} \mathrm{CuCl}_{2}$. For 10 days of cultivation, bacteria of all strains didn't completely reduce the copper (II) ions present in the medium.

To study the influence of $\mathrm{Cu}^{2+}$ on metal reducing activity of Desulfuromonas sp., bacteria cells, previously grown in a medium with fumarate were incubated for 1 hour with $0.5-4.0 \mathrm{mM} \mathrm{CuCl}_{2}$, washed and cultivated in media with cysteine as sulphur source and $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}$, $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}$ (control), $\mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ at concentrations of 1.74, 3.47, $5.21,6.94,10.41 \mathrm{mM}$. It has been established that with increasing concentrations of ferrum (III), chromium (VI) or manganese (IV) compounds in the medium, biomass accumulation by Desulfuromonas sp. non-incubated and incubated with $\mathrm{CuCl}_{2}$ cells after 10 days of growth decreased (Fig. 3, 4).



Fig. 3. The influence of $\mathrm{CuCl}_{2}$ on biomass accumulation by $D$. acetoxidans IMV B-7384 (a, $d$ ), Desulfuromonas sp. Yavor-5 ( $b, e$ ) and Desulfuromonas sp. Yavor-7 $(c, f)$ after 10 days of growth in medium with $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}(a, b, c)$ or $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}(d, e, f)$ : ${ }^{*}-\mathrm{P}<0.05$


Fig. 4. The influence of $\mathrm{CuCl}_{2}$ on biomass accumulation by D. acetoxidans IMV B-7384 (a,d), Desulfuromonas sp. Yavor-5 ( $b, e$ ) and Desulfuromonas sp. Yavor-7 $(c, f)$ after 10 days of growth in medium with $\mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}(a, b, c)$ or $\mathrm{MnO}_{2}(d, e, f): *-\mathrm{P}<0.05$

In contrast to the growth of bacteria in media with metal compounds, with increasing of fumarate concentrations in the cultivation medium $1.74-5.21 \mathrm{mM}$, a rise in biomass accumulation by bacteria of all strains was observed, which decreased with further growth of its concentrations up to 10.41 mM . The bacteria cells, not incubated with $\mathrm{CuCl}_{2}$, for 10 days accumulated the highest biomass in the medium with ferric (III) citrate (up to $2.77 \mathrm{~g} / \mathrm{l}$ ), in comparison with growth in media with fumarate (up to $2.60 \mathrm{~g} / \mathrm{l}$ ), potassium dichromate (up to $1.52 \mathrm{~g} / \mathrm{l}$ ) or manganese (IV) oxide (up to $1.37 \mathrm{~g} / \mathrm{l}$ ). With increasing concentrations of $\mathrm{CuCl}_{2}$ during incubation, the biomass accumulation by bacteria in media with all tested electron acceptors decreased, which can be explained by the toxic effect of copper (II) ions on bacteria cells. However, copper (II) chloride at concentrations of $0.5-2.5 \mathrm{mM}$ in the incubation mixture didn't have a significant negative effect on bacteria growth in media, both with fumarate and with metal compounds at all concentrations. Under the influence of $3.0-4.0 \mathrm{mM} \mathrm{CuCl} 2$, the biomass of bacteria in media with $1.74-10.41 \mathrm{mM} \mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}$ and $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}$ decreased by 2.1-2.9 and 2.6-4.9 times, respectively (Fig. 3). Under the
influence of $2.5 \mathrm{mM} \mathrm{CuCl}_{2}$, the biomass accumulation by bacteria in media with $1.74-10.41 \mathrm{mM} \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ and $\mathrm{MnO}_{2}$ decreased by 2.4 and 2.6 times respectively, and incubated with $3.0-4.0 \mathrm{mM} \mathrm{CuCl}_{2}$ cells of all strains in media with these electron acceptors practically didn't grow, probably due to the negative effect on them of potassium dichromate or manganese (IV) oxide and $\mathrm{Cu}^{2+}$ (Fig. 4).

Thus, it was found that metal reducing activity of sulphur-reducing bacteria cells, incubated with $0.5-2.0 \mathrm{mM} \mathrm{CuCl}_{2}$, remained almost unchanged compared to activity of non-incubated cells, which demonstrates the high potential of application of these microorganisms in biotechnologies for purification of environments with complex pollution from heavy metals. Copper (II) chloride at concentrations of $2.5-3.0 \mathrm{mM}$ in the incubation mixture inhibited growth of all Desulfuromonas sp. strains in media with $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ as electron acceptors at concentrations $1.74-10.41 \mathrm{mM}$.

## Discussion

Hydrogen sulfide, produced by sulfidogenic microbiota, interacts with heavy metal ions, precipitating them in form of sulfides (White
et al., 2000; Kiran et al., 2017). The formation of metal sulpfides is the main way of eliminating heavy metals from the natural cycle (White et al., 2000; Kozlova et al., 2008; Kuznetsov et al., 2015). The efficiency of microbiological precipitation of metal ions by hydrogen sulfide, produced by cells of sulphur-reducing bacteria, depends on concentration of $\mathrm{H}_{2} \mathrm{~S}$, which they produced in process of dissimilatory sulphur reduction (Gudz et al., 2011). On the other hand, the intensity of anaerobic respiration in these microorganisms in contaminated ecotopes is determined by level of their adaptation to unfavourable environmental conditions, in particular, increased content of metal compounds (Viti et al., 2014; Si et al., 2015; Simonte et al., 2017). Among the bacteria of the Desulfuromonas genus, no resistance to copper species, or those that carry out its reduction, are described (Vandieken et al., 2006).

Table 1
Copper (II) sulfide formation in sulphur-reducing bacteria cultivation medium after addition $\mathrm{S}^{0}$ and $\mathrm{CuCl}_{2}$

| Strain | $\mathrm{CuCl}_{2}, \mathrm{mM}$ | $\mathrm{CuS}, \mathrm{mM}$ | Copper (II) ions <br> binding, $\%$ | Qualitative analysis <br> on ions presence |
| :---: | :---: | :---: | :---: | :---: |
|  | control | 0 | 0 | - |
|  | 0.5 | $0.50 \pm 0.05$ | $100.0 \pm 0.2$ | - |
| D. acetoxidans | 1.0 | $0.98 \pm 0.02$ | $98.0 \pm 0.3$ | - |
| IMV B-7384 | 1.5 | $1.46 \pm 0.01$ | $97.3 \pm 0.1$ | - |
|  | 2.5 | $1.58 \pm 0.04$ | $79.0 \pm 0.4$ | + |
|  | 3.0 | $1.50 \pm 0.03$ | $60.0 \pm 0.5$ | + |
|  | 4.0 | $0.94 \pm 0.01$ | $39.7 \pm 0.1$ | + |
|  | control | 0 | $23.5 \pm 0.2$ | + |
|  | 0.5 | $0.50 \pm 0.03$ | $100.0 \pm 0.1$ | - |
| Desulfuro- | 1.0 | $1.00 \pm 0.01$ | $100.0 \pm 0.4$ | - |
| monas sp. | 1.5 | $1.47 \pm 0.03$ | $98.0 \pm 0.3$ | - |
| Yavor-5 | 2.0 | $1.60 \pm 0.02$ | $80.0 \pm 0.5$ | - |
|  | 2.5 | $1.52 \pm 0.04$ | $60.8 \pm 0.1$ | + |
|  | 3.0 | $1.21 \pm 0.05$ | $40.3 \pm 0.2$ | + |
|  | 4.0 | $0.95 \pm 0.01$ | $23.8 \pm 0.1$ | + |
|  | control | 0 | 0 | + |
|  | 0.5 | $0.50 \pm 0.02$ | $100.0 \pm 0.1$ | - |
| Desulfuro- | 1.0 | $1.00 \pm 0.03$ | $100.0 \pm 0.4$ | - |
| monas sp. | 1.5 | $1.49 \pm 0.04$ | $99.3 \pm 0.3$ | - |
| Yavor-7 | 2.0 | $1.62 \pm 0.02$ | $81.0 \pm 0.2$ | - |
|  | 2.5 | $1.53 \pm 0.02$ | $61.2 \pm 0.1$ | + |
|  | 3.0 | $1.22 \pm 0.01$ | $40.7 \pm 0.4$ | + |
|  | 4.0 | $0.96 \pm 0.03$ | $24.0 \pm 0.5$ | + |

Note: "+" - copper (II) ions are present; "-" - copper (II) ions are absent.
To study the influence of copper (II) ions on sulfidogenic activity of sulphur-reducing bacteria, cells were incubated for one hour with $\mathrm{CuCl}_{2}$ at different concentrations, washed and cultivated in a medium with elemental sulphur to avoid, on the one hand, extracellular reduction of $\mathrm{Cu}^{2+}$, and, on the other, formation in the cultural liquid of insoluble metal sulfide. Often, it takes just five minutes to completely absorb metals with bacteria, but this time can range from a few seconds to an hour after contact with toxicant (White et al., 2000; Tashirev et al., 1995). It was assumed that the inhibition of metabolic processes in bacteria incubated with copper (II) ions would be due to their accumulation by surface structures (as result of binding with polysaccharides, outer membrane proteins, cell wall structures) or inside cells, where they may not inversely interact with structural components of cytoplasmic membrane, cytoplasm proteins and other metabolites (Lengeler et al., 2005; Tashirev et al., 2008). To determine the level of copper (II) ions precipitation by hydrogen sulfide produced by bacteria, cells were cultivated in a medium with $\mathrm{S}^{0}$, to which $\mathrm{CuCl}_{2}$ was added at different concentrations. Despite that, in case of simultaneous presence in medium $\mathrm{S}^{0}$ and oxidized form of copper, bacteria will use (as assumed) $\mathrm{CuCl}_{2}$ as the more energy efficient electron acceptor (oxidation-reduction potential of pair $\mathrm{Cu}^{2+} / \mathrm{Cu}^{+} E_{0}{ }^{\prime}=+0.15 \mathrm{~V}$ is higher than that of pair $\mathrm{S}^{0} / \mathrm{H}^{\mathrm{S}} E_{0}{ }^{\prime}=$ -0.27 V) (Sani et al., 2001; Lengeler et al., 2005; Richter et al., 2012), at high concentrations it showed a more or less pronounced toxic effect on the microorganisms.

Previously, we found that bacteria of the Desulfuromonas genus, isolated from reservoir of the Yavoriv sulphur deposit, in process of anaerobic destruction of organic compounds with different intensity use
up to $10 \mathrm{mM} \mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ and $\mathrm{MnO}_{2}$ as final electron acceptors (Moroz et al., 2014, 2017). Therefore we studied the ability of these bacteria to reduce copper (II) cations in the process of anaerobic respiration. Whereas reduction of metals pro-oxidants by membrane-bound metal reductases is carried out outside the cell (Gescher \& Kappler, 2012; Richter et al., 2012; Simonte et al., 2017), with increasing concentration of soluble metal compounds in the medium, the degree of their cations or oxoanions penetration through bacteria cytoplasmic membrane in cytoplasm increases, where their interaction with intracellular metabolites occurs, chemically active intermediates, oxygen radicals are formed, reduced forms of metals as end products are accumulated, which causes the inhibitory effect on growth (Richter et al., 2012; Viti et al., 2014).

Table 2
$\mathrm{Cu}^{2+}$ reducing by sulphur-reducing bacteria after addition into medium $\mathrm{CuCl}_{2}$

| Strain | Cultivation time, days | $\mathrm{Cu}^{2+}, \mathrm{mM}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1.74 | 3.47 | 5.21 | 6.94 | 10.41 |
| $\begin{gathered} \hline \text { D. acetoxidans IMV } \\ \text { B-7384 } \end{gathered}$ | 0 | + | + | + | $+$ | + |
|  | 2 | + | + | + | + | + |
|  | 4 | $+$ | + | $+$ | $+$ | $+$ |
|  | 6 | + | + | $+$ | $+$ | $+$ |
|  | 8 | + | $+$ | $+$ | $+$ | $+$ |
|  | 10 | + | + | + | + | + |
| Desulfuromonas sp. Yavor-5 | 0 | + | + | + | $+$ | + |
|  | 2 | $+$ | $+$ | $+$ | $+$ | $+$ |
|  | 4 | + | + | + | + | + |
|  | 6 | + | $+$ | $+$ | $+$ | $+$ |
|  | 8 | + | + | + | + | + |
|  | 10 | + | + | $+$ | $+$ | + |
| Desulfuromonas sp. Yavor-7 | 0 | + | + | $+$ | $+$ | + |
|  | 2 | $+$ | + | $+$ | $+$ | $+$ |
|  | 4 | $+$ | + | $+$ | $+$ | $+$ |
|  | 6 | $+$ | + | $+$ | $+$ | $+$ |
|  | 8 | $+$ | + | $+$ | $+$ | $+$ |
|  | 10 | + | + | + | $+$ | + |

Note: "+" - copper (II) ions are present; "-" - copper (II) ions are absent.
The efficiency of biological methods for purifying the environment from pollutants depends not only on metabolic activity of selected strains of bacteria, but primarily on their resistance to metal ions. Therefore, the influence of $\mathrm{CuCl}_{2}$ at concentrations of $0.5,1.0,1.5,2.0$, 3.0 times different from the sulphur content in Kravtsov-Sorokin medium of standard composition on biomass accumulation by sulphurreducing bacteria was investigated. It is known that bacteria reduce fumarate to succinate in the process of fumarate respiration with participation of electron transport chain, which contains a number of dehydrogenases and fumarate reductase, which are interconnected by pool of $b$-type cytochromes and menaquinones (Lengeler et al., 2005). In a medium with $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}, \mathrm{H}_{2} \mathrm{~S}$ toxic for cells does not form and biomass yield is almost the same as in the medium with $\mathrm{S}^{0}$ (Moroz et al., 2017), so a medium with fumarate was used as control to compare the biomass accumulation by bacteria during usage of copper (II) chloride or other metals compounds and $\mathrm{C}_{4} \mathrm{H}_{4} \mathrm{O}_{4}$ as electron acceptors. Although oxida-tion-reduction potential of the pair $\mathrm{Cu}^{2+} / \mathrm{Cu}^{+}\left(E_{0}{ }^{\prime}=+0.15 \mathrm{~V}\right)$ is higher than that of the pair fumarate/succinate $\left(E_{0}{ }^{\prime}=+0.03 \mathrm{~V}\right)$ (Lengeler et al., 2005), the use of $\mathrm{CuCl}_{2}$ by microorganisms was very slow.

To study the influence of copper (II) ions on metal reducing activity of bacteria, the cells were incubated for one hour with $\mathrm{CuCl}_{2}$ at different concentrations, washed (to avoid extracellular reduction of $\mathrm{Cu}^{2+}$ ) and cultivated in media with ferric (III) citrate, potassium dichromate, manganese (IV) oxide or fumarate (control) at concentrations $1.74-10.41 \mathrm{mM}$. With increasing $\mathrm{CuCl}_{2}$ concentrations during incubation, the growth of bacteria in media with all electron acceptors decreased, possibly due to its toxic effect on bacteria cells. Although oxidation-reduction potential of the pair $\mathrm{Cr}_{2} \mathrm{O}_{7}{ }^{2-} / \mathrm{Cr}^{3+}\left(E_{0}{ }^{\prime}=+1.33 \mathrm{~V}\right)$ is higher than that of the pairs Mn (IV) $/ \mathrm{Mn}$ (II) $\left(E_{0}{ }^{\prime}=+1.23 \mathrm{~V}\right)$ and Fe (III) $/ \mathrm{Fe}$ (II) $\left(E_{0}{ }^{\prime}=+0.77 \mathrm{~V}\right)$ and much higher than that of the pair fumarate/succinate $\left(E_{0}{ }^{\prime}=+0.03 \mathrm{~V}\right)$
(Lengeler et al., 2005), use of potassium dichromate and manganese (IV) oxide by microorganisms did not provide the high biomass accumulation.

Selection of strains of microorganisms from anthropogenically altered biotopes, which are characterized by increased content of sulphur and metal compounds, deserves special attention. Sulphur-reducing bacteria D. acetoxidans IMV B-7384, Desulfuromonas sp. Yavor-5 and Desulfuromonas sp. Yavor-7, isolated by us from Yavorivske Lake, were revealed to be highly resistant to copper (II) ions at concentrations much higher than those in this technogenic reservoir. Studies of their sulphurand metal-reducing activities under the influence of stress factors, in particular action of $\mathrm{CuCl}_{2}$ and compounds of other metals, is especially important, since formation of metal sulfides is the main way by which heavy metals are removed from the natural cycle (White et al., 2000; Kuznetsov et al., 2015; Kiran et al., 2017), and reduction of transitional heavy metal compounds by microorganisms in media with low oxida-tion-reduction potential plays an important role in the process of organic substrates oxidation (Kozlova et al., 2008; Gescher \& Kappler, 2012; Simonte et al., 2017). Almost complete precipitation up to 1.5 mM of copper (II) ions in form of CuS by $\mathrm{H}_{2} \mathrm{~S}$ produced by bacteria and ability to reduce up to $10.41 \mathrm{mM} \mathrm{CuCl}_{2}, \mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ in the process of anaerobic respiration indicates a high adaptation of the bacteria strains investigated by us to stress factors, in particular, the influence of copper (II) chloride.

Investigation of mechanisms of interaction of sulfidogenic representatives of microbiota from Yavorivske Lake with heavy metals prooxidants which are highly toxic to microorganisms, such as $\mathrm{Cu}^{2+}$, are necessary for the widest possible disclosure of their potential in transformation of sulphur, carbon and metal compounds and creation on their basis of new methods for protecting the environment from hazardous chemical pollutants.

## Conclusions

Thus, it was established that the sulfidogenic activity of bacteria strains Desulfuromonas sp., isolated from Yavorivske Lake, under the influence of $0.5-2.5 \mathrm{mM} \mathrm{CuCl}_{2}$ practically did not change. The efficiency of copper (II) ions precipitation in the form of CuS by hydrogen sulfide produced by bacteria reached $100.0 \%$ after addition into a medium with sulphur of up to $1.5 \mathrm{mM} \mathrm{CuCl}_{2}$. Bacteria used copper (II) ions as an electron acceptor in the process of anaerobic respiration at concentrations of $1.74-10.41 \mathrm{mM}$ in a medium with their incomplete reduction for 10 days of cultivation. The copper (II) chloride at concentrations of 0.52.0 mM in the incubation mixture did not significantly affect the biomass accumulation by bacteria in media with $1.74-10.41 \mathrm{mM}$ $\mathrm{C}_{6} \mathrm{H}_{5} \mathrm{O}_{7} \mathrm{Fe}, \mathrm{K}_{2} \mathrm{Cr}_{2} \mathrm{O}_{7}$ or $\mathrm{MnO}_{2}$ as electron acceptors, which demonstrates the high application potential of these microorganisms in biotechnologies for purifying environments with complex pollution from heavy metals.

## References

Behlau, F., Gochez, A. M., Lugo, A. J., Elibox, W., Minsavage, G. V., Potnis, N., White, F. F., Ebrahim, M., Jones, J. B., \& Ramsubhag, A. (2017). Characterization of a unique copper resistance gene cluster in Xanthomonas campestris pv. campestris isolated in Trinidad, West Indies. European Journal of Plant Pathology, 147(3), 671-681.
Betehtin, A. G. (2007). Kurs mineralogii [The course of mineralogy]. KDU, Moscow (in Russian).
Bilyy, O. I. Vasyliv, O. M., \& Hnatush, S. O. (2014). The anode biocatalyst with simultaneous transition metals pollution control. Technology and Application of Microbial Fuel Cells. InTech, Rijeka, Croatia.
Breuer, M., Rosso, K. M., Blumberger, J., \& Butt, J. N. (2015). Multi-haem cytochromes in Shewanella oneidensis MR-1: Structures, functions and opportunities. Journal of the Royal Society Interface, 12(102), 20141117.
Cidre, I., Pulido, R. P., Grande Burgos, M. J., Gálvez, A., \& Lucas, R. (2017). Copper and zinc tolerance in bacteria isolated from fresh produce. Journal of Food Protection, 80(6), 969-975.
Dey, U., Chatterjee, S., \& Mondal, N. K. (2016). Isolation and characterization of arsenic-resistant bacteria and possible application in bioremediation. Biotechnology Reports, 10, 1-7.
Diakiv, S. V., Hnatush, S. O., Babenko, V. V., \& Moroz, O. M. (2017). Sul'fidogenna aktyvnist' Desulfuromusa sp. SV30 za vplyvu spoluk hromu, kuprumu
i ferumu [Sulfidogenic activity of Desulfuromusa sp. SV30 under the influence of chromium, copper and iron compounds]. Studia Biologica, 11(2), 53-66 (in Ukrainian).
Diakiv, S. V., Hnatush, S. O., Moroz, O. M., Prypin, O. Y., Kulachkovskyi, O. R., \& Bodnaruk, V. Y. (2016). Sulfur reducing bacteria from coal pits waste heaps of Chervonograd Mining Region. Studia Biologica, 10(2), 63-76.
Frank, Y. A., \& Lushnikov, S. V. (2006). Biotekhnologicheskij potencial sul'fatreduciruyushchih bakterij [Biotechnological potential of sulfate reducing bacteria]. Ehkologiya i Promyshlennost', 1, 10-13 (in Russian).
Gescher, J., \& Kappler, A. (2012). Microbial metal respiration: From geochemistry to potential applications. Springer-Verlag, Heidelberg, Berlin.
Gillan, D. C., van Camp, C., Mergeay, M., Provoost, A., Thomas, N., Vermard, L., Billon, G., \& Wattiez, R. (2017). Paleomicrobiology to investigate copper resistance in bacteria: Isolation and description of Cupriavidus necator B9 in the soil of a medieval foundry. Environmental Microbiology, 19(2), 770-787.
Gochez, A. M., Huguet-Tapia, J. C., Minsavage, G. V., Shantaraj, D., Jalan, N., Strauß, A., Lahaye, T., Wang, N., Canteros, B. I., Jones, J. B., \& Potnis, N. (2018). PacBio sequencing of copper-tolerant Xanthomonas citri reveals presence of a chimeric plasmid structure and provides insights into reassortment and shuffling of transcription activator-like effectors among $X$. citri strains. BMC Genomics, 19(16), 1-14.
Gudz, S. P., Peretiatko, T. B., Moroz, O. M., Hnatush, S. O., \& Klym, I. R. (2011). Rehulyuvannya rivnya sul'fativ, sirkovodnyu ta vazhkykh metaliv u tekhnohennykh vodoymakh sulfatvidnovljuval'nymy bakterijamy [Regulation of sulfates, hydrogen sulfide and hard metals level in technogenic reservoirs by sulfate reducing bacteria]. Mikrobiologichny Zhurnal, 73(2), 33-38 (in Ukrainian).
Gudz, S. P., Hnatush, S. O., Moroz, O. M., Peretiatko, T. B., \& Vasyliv, O. M. (2013) Svidotstvo pro deponuvannya shtamu bakteriy Desulfuromonas acetoxidans Ya-2006 u Depozytariyi Instytutu mikrobiolohiyi i virusolohiyi im. D. K. Zabolotnoho NAN Ukrayiny z nadannyam reyestratsiynoho nomeru IMV B7384 [Certificate of deposition of bacteria Desulfuromonas acetoxidans Ya2006 strain at the Depository of D. K. Zabolotny Institute of Microbiology and Virology of the NAS of Ukraine with appropriation of registration number IMV B-7384] (in Ukrainian).
Gudz, S. P., Hnatush, S. O., Yavorska, G. V., Bilinska, I. S., \& Borsukevych, B. M. (2014). Praktykum z mikrobiologii' [Workshop on microbiology]. Ivan Franko National University of L'viv, Lviv (in Ukrainian).
Harris, D. S. (2003). Quantitative chemical analysis. Amazon, New York.
Hedderich, R., Klimmek, O., Kroger, A., Dirmeier, R., Keller, M., \& Stetter, K. O. (1999). Anaerobic respiration with elemental sulfur and with disulfides. FEMS Microbiology Reviews, 22(5), 353-381.
Iwahori, K., Watanabe, J., Tani, Y., Seyama, H., \& Miyata, N. (2014). Removal of heavy metal cations by biogenic magnetite nanoparticles produced in Fe(III)-reducing microbial enrichment cultures. Journal of Bioscience and Bioengineering, 117(3), 333-335.
Kiran, M. G., Pakshirajan, K., \& Das, G. (2017). Heavy metal removal from multicomponent system by sulfate reducing bacteria: Mechanism and cell surface characterization. Journal of Hazardous Materials, 324(A), 62-70.
Kozlova, I. P., Radchenko, O. S., Stepura, L. H., Kondratyuk, T. O., \& Pilyashen-ko-Novokhatnyy, A. I. (2008). Heokhimichna diyalnist mikroorhanizmiv ta yiyi prykladni aspekty [Geochemical activity of microorganisms and its applied aspects]. Naukova Dumka, Kyiv (in Ukrainian).
Kucherjavyj, V. P. (2001). Ekologija [Ecology]. Svit, Lviv (in Ukrainian)
Kumar, A., Bisht, B. S., \& Joshi, V. D. (2010). Biosorption of heavy metals by four acclimated species, Bacillus spp., Pseudomonas spp., Staphylococcus spp. and Aspergillus niger. Journal of Biological and Environmental Sciences, 4(12), 97-108.
Kuznetsov, A., Gradova, N., Lushnikov, S., Éngelkhart, M., Vaysser, T., \& Chebotareva, M. (2015). Prikladnaya ehkobiotekhnologiya [Applied Ecobiotechnology]. Binom Laboratoriya Znanij, Moscow (in Russian).
Ladomersky, E., \& Petris, M. J. (2015). Copper tolerance and virulence in bacteria. Metallomics, 7, 957-964.
Lengeler, J., Drevs, G., \& Shlegel, G. (Eds.). (2005). Sovremennaya mikrobiologiya. Prokarioty [Contemporary Microbiology. Prokaryotes]. Mir, Moscow (in Russian).
Limcharoensuk, T., Sooksawat, N., Sumarnrote, A., Awutpet, T., Kruatrachue, M., Pokethitiyook, P., \& Auesukaree, C. (2015). Bioaccumulation and biosorption of $\mathrm{Cd}^{2+}$ and $\mathrm{Zn}^{2+}$ by bacteria isolated from a zinc mine in Thailand. Ecotoxicology and Environmental Safety, 122, 322-330.
Martínez-Bussenius, C., Navarro, C. A., \& Jerez, C. A. (2017). Microbial copper resistance: Importance in biohydrometallurgy. Microbial Biotechnology, 10(2), 279-295.
Maslovska, O., \& Hnatush, S. (2015). Oxidative modification of proteins and specific superoxide dismuase activity of Desulfuromonas acetoxidans IMV B-7384 bacteria under the influence of ferric citrate. Microbiology and Biotechnology, 30, 34-40.

Moroz, O. M. (2013). Utvorennya hidrohen sul'fidu sirkovidnovlyuval'nymy bakteriyamy za vplyvu soley vazhkykh metaliv [Hydrogen sulfide production by sulfur reducing bacteria under the influence of heavy metal salts]. Visnyk of L'viv University. Biological Series, 61, 154-165 (in Ukrainian)
Moroz, O. M., Hnatush, S. O., Bohoslavets, C. I., Yavorska, G. V., Zvir, G. I., \& Borsukevych, B. M. (2017). Vplyv kalij bihromatu na dejaki fiziologichni osoblyvosti bakterij cyklu sul'furu ozera Javorivs'ke [Potassium dichromate influence on some physiological peculiarities of sulfur cycle bacteria from Yavorivske lake]. Visnyk of L'viv University. Biological Series, 75, 127139 (in Ukrainian).
Moroz, O., Gul', N., Galushka, A., Zvir, G., \& Borsukevych, B. (2014). Vykorystannja riznyh akceptoriv elektroniv bakterijamy Desulfuromonas sp., vydilenymy z ozera Javorivs'ke [Different electron acceptors usage by bacteria of Desulfuromonas sp. isolated from Yavorivske lake]. Visnyk of L'viv University. Biological Series, 65, 322-334 (in Ukrainian).
Mustapha, M. U., \& Halimoon, N. (2015). Screening and isolation of heavy metal tolerant bacteria in industrial effluent. Procedia Environmental Sciences, 30, 33-37.
Richard, D., Boyer, C., Vernière, C., Canteros, B. I., Lefeuvre, P., \& Pruvost, O. (2017a). Complete genome sequences of six copper-resistant Xanthomonas citri pv. citri strains causing Asiatic citrus canker, obtained using long-read technology. Genome Announcements, 5(12), e00010-17.
Richard, D., Ravigné, V., Rieux, A., Facon, B., Boyer, C., Boyer, K., Grygiel, P., Javegny, S., Terville, M., Canteros, B. I., Robène, I., Vernière, C., Chabirand, A., Pruvost, O., \& Lefeuvre, P. (2017b). Adaptation of genetically monomorphic bacteria: Evolution of copper resistance through multiple horizontal gene transfers of complex and versatile mobile genetic elements. Molecular Ecology, 26(7), 2131-2149.
Richter, K., Schicklberger, M., \& Gescher, J. (2012). Dissimilatory reduction of extracellular electron acceptors in anaerobic respiration. Applied Environmental Microbiology, 78(4), 913-921.
Sani, R. K., Peyton, B. M., \& Brown, L. T. (2001). Copper-induced inhibition of growth of Desulfovibrio desulfuricans G20: Assessment of its toxicity and correlation with those of zinc and lead. Applied and Environmental Microbiology, 67(10), 4765-4772.
Segin, T., Hnatush, S., \& Gorishniy, M. (2016). Protsesy lipoperoksydatsiyi u klitynakh Chlorobium limicola IMV K-8 za vplyvu Cu (II) sul'fatu [Lipoperoxidation processes in Chlorobium limicola IMV K-8 cells under the
influence of Cu (II) sulfate]. Visnyk of Dnipropetrovsk University. Biology, Ecology, 24(1), 72-78 (in Ukrainian).
Si, Y., Zou, Y., Liu, X., Si, X., \& Mao, J. (2015). Mercury methylation coupled to iron reduction by dissimilatory iron-reducing bacteria. Chemosphere, 122, 206-212.
Silver, S., \& Walderhaug, M. (1995). Bacterial plasmid-mediated resistances to mercury, cadmium and copper. Toxicology of metals. Biochemical aspects. Springer, Berlin.
Simonte, F., Sturm, G., Gescher, J., \& Sturm-Richter, K. (2017). Extracellular electron transfer and biosensors. Advances in Biochemical Engineering / Biotechnology. Springer, Berlin, Heidelberg.
Solioz, M., \& Stoyanov, J. (2003). Copper homeostasis in Enterococcus hirae. FEMS Microbiology Reviews, 27(2-3), 183-195.
Tashirev, A. B. (1995). Vzaimodejstvie mikroorganizmov s metallami [Interaction of microorganisms with metals]. Mikrobiologicheskij Zhurnal, 57(2), 95-104 (in Russian).
Tashirev, A. B., Galinker, J. V., \& Andrejuk, E. I. (2008). Termodinamicheskoe prognozirovanie redoks-vzaimodejstvija mikroorganizmov s metallami-okisliteljami $\left(\mathrm{Hg}^{2+}, \mathrm{CrO}_{4}{ }^{2-} \mathrm{iCu}^{2+}\right)$ [Thermodynamic calculations of redox interaction of microorganisms with oxidative metals $\left(\mathrm{Hg}^{2+}, \mathrm{CrO}_{4}{ }^{2-}\right.$ and $\left.\left.\mathrm{Cu}^{2+}\right)\right]$. Doklady NAN Ukrainy, 4, 166-172 (in Russian).
Tashirev, A. B., Matveeva, N. A., Romanovskaja, V. A., Tashireva, A. A., \& Rokitko, P. V. (2007). Polirezistentnost' i sverhustojchivost' $k$ tjazhjolym metallam antarkticheskih mikroorganizmov [Polyresistance and superstability of Antarctic microorganisms to heavy metals]. Doklady NAN Ukrainy, 11, 170-175 (in Russian).
Vandieken, V., Mussmann, M., Niemann, H., \& Jørgensen, B. B. (2006). Desulfuromonas svalbardensis sp. nov. and Desulfuromusa ferrireducens sp. nov., psychrophilic, Fe(III)-reducing bacteria isolated from arctic sediments, Svalbard. International Journal of Systematic and Evolutionary Microbiology, 56, 1133-1139.
Viti, C., Marchi, E., Decorosi, F., \& Giovannetti, L. (2014). Molecular mechanisms of Cr (VI) resistance in bacteria and fungi. FEMS Microbiology Reviews, 38(4), 633-659.
White, C., Sayer, J. A., \& Gadd, G. M. (2000). Microbial solubilization and immobilization of toxic metals: Key biogeochemical processes for treatment of contamination. FEMS Microbiology Ecology, 33, 197-208.
Winkelmann, G. (Ed.). (2008). Microbial transport systems. Wiley-VCH, New York.

