



Article

# Sequestering Atmospheric CO<sub>2</sub> Inorganically: A Solution for Malaysia's CO<sub>2</sub> Emission

M. Ehsan Jorat 1,\*, Maniruzzaman A. Aziz 2, Aminaton Marto 3, Nabilah Zaini 4, Siti Norafida Jusoh <sup>5</sup> and David A.C. Manning <sup>6</sup>

- Division of Natural and Built Environment, School of Science Engineering and Technology, Abertay University, Dundee DD1 1HG, UK
- Department of Geotechnics and Transportation, School of Civil Engineering, Faculty of Engineering, Universiti Teknologi Malaysia, 81310 Johor Bahru, Johor, Malaysia; mzaman@utm.my
- Department of Environmental Engineering and Green Technology, Malaysia-Japan International Institute of Technology, Universiti Teknologi Malaysia, 54100 Kuala Lumpur, Malaysia; aminaton@utm.my
- Department of Chemical Process Engineering, Malaysia-Japan International Institute of Technology, Universiti Teknologi Malaysia, 54100 Kuala Lumpur, Malaysia; nabilah.zaini@utm.my
- Department of Geotechnics and Transportation, School of Civil Engineering, Faculty of Engineering, Universiti Teknologi Malaysia, 81310 Johor Bahru, Johor, Malaysia; snorafida@utm.my
- School of Natural and Environmental Sciences, Newcastle University, Newcastle upon Tyne, NE1 7RU, UK; david.manning@newcastle.ac.uk
- Correspondence: e.jorat@abertay.ac.uk; Tel.: +44-1382-30-8930

Received: 13 November 2018; Accepted: 11 December 2018; Published: 14 December 2018



**Abstract:** Malaysia is anticipating an increase of 68.86% in CO<sub>2</sub> emission in 2020, compared with the 2000 baseline, reaching 285.73 million tonnes. A major contributor to Malaysia's CO<sub>2</sub> emissions is coal-fired electricity power plants, responsible for 43.4% of the overall emissions. Malaysia's forest soil offers organic sequestration of 15 tonnes of  $CO_2$  ha<sup>-1</sup>·year<sup>-1</sup>. Unlike organic  $CO_2$  sequestration in soil, inorganic sequestration of CO<sub>2</sub> through mineral carbonation, once formed, is considered as a permanent sink. Inorganic CO<sub>2</sub> sequestration in Malaysia has not been extensively studied, and the country's potential for using the technique for atmospheric CO<sub>2</sub> removal is undefined. In addition, Malaysia produces a significant amount of solid waste annually and, of that, demolition concrete waste, basalt quarry fine, and fly and bottom ashes are calcium-rich materials suitable for inorganic CO<sub>2</sub> sequestration. This project introduces a potential solution for sequestering atmospheric CO<sub>2</sub> inorganically for Malaysia. If lands associated to future developments in Malaysia are designed for inorganic CO<sub>2</sub> sequestration using demolition concrete waste, basalt quarry fine, and fly and bottom ashes, 597,465 tonnes of CO<sub>2</sub> can be captured annually adding a potential annual economic benefit of €4,700,000.

Keywords: CO<sub>2</sub> emission; Malaysia; inorganic CO<sub>2</sub> sequestration; demolition concrete waste; basalt quarry fine; fly and bottom ash

# 1. Introduction

According to the International Energy Agency [1], the world's annual carbon dioxide (CO<sub>2</sub>) emissions grew from 17.78 billion tonnes in 1980, to 32.10 billion tonnes in 2015. Stern [2] reported that, if no action is taken to reduce greenhouse gas (GHG) emissions, the concentration of GHGs in the atmosphere would become double its preindustrial level by as early as 2035. According to the Institute of Energy Economics, Japan (IEEJ) 2018 Outlook—Prospects and Challenges until 2050 [3], the global energy demand continues to increase from 1990 to 2050. Based on this predicted energy outlook, two thirds of the energy growth comes from non-Organisation for Economic Co-operation and

Development (OECD) Asia, including China, India, Association of Southeast Asian Nations (ASEAN) countries, and others in Asia, Middle East, North Africa, sub-Saharan Africa, Europe, and Latin America. Three quarters of the energy growth until 2050 will be utilised for fueling power generation and transportation. Coal, natural gas, and biomass are the three main sources for electricity production. The International Energy Agency [4] indicated that the electricity produced using fossil fuels is responsible for nearly 40% of the world's overall energy-related CO<sub>2</sub> emissions. CO<sub>2</sub> emissions related to the main carbon emitting industries such as iron and steel industry, cement industry, petroleum refining, and the pulp and paper industry accounts for 30% of the total global anthropogenic emission of CO<sub>2</sub>, responsible for a significant portion of industry-related CO<sub>2</sub> emissions [5]. However, emissions related to power generation still stand as the main contributor to industry-related CO<sub>2</sub> emissions. According to the statistical data from Global Report on Human Settlement, Source on the Cities and climate change, about 75% of CO<sub>2</sub> increment in atmosphere originates from industrial power plants, especially from the burning of fossil fuels [6].

Malaysia is anticipating an increase of 69% in  $CO_2$  emissions in 2020, compared with 2000 baseline, reaching 285.7 million tonnes [7]. Considering Malaysia's population, the per capita GHG emission is 5.9 million tonnes which is three times higher than the levels anticipated for the whole Southeast Asia region [8]. The major contribution to the country's  $CO_2$  emissions is from coal-fired electricity power plants, which are responsible for 43.4% of the overall emissions. The total coal consumption for electricity generation in Malaysia is projected to increase from 12.4 million tonnes in 2005 to 36 million tonnes in 2020 [9].

In 2016 Earth Day, 174 countries singed the 2015 United Nations Climate Change Conference (COP21) Agreement to limit the global warming to 2  $^{\circ}$ C based on the pre-industrial level. Malaysia is one of the 197 countries that signed the Agreement for active participation in mitigating the carbon emissions by 2030 [10]. The Paris Agreement has acknowledged the requirement to (i) reduce  $CO_2$  emissions and (ii) remove existing  $CO_2$  from atmosphere. Carbon capture and storage (CCS) had been acknowledged as a process of removing  $CO_2$  at high concentration for the long-term duration. This strategy has high potential in mitigating 7 to 70% of cumulative mitigation effort globally by 2100 [11].

Malaysia is committed to the COP21 Paris Agreement. The major reduction in CO<sub>2</sub> emission is likely to be accomplished through the improved efficiency in the energy systems (e.g., low carbon energy and fuels), but the implementation of CCS development programmes would make a significant contribution) to meet the agreement's objectives. Malaysia is one of the Southeast Asia countries investing on minimising concentration of greenhouse gases (GHGs) in general and CO<sub>2</sub> through defined strategies falling under CCS development [12].

 $CO_2$  removal in geological and terrestrial reservoirs through anthropogenic activities are recommended as a method to fight global warming [13]. Mineral carbonation and using soil as a sink for carbon has been proven as an effective CCS method for permanent atmospheric  $CO_2$  removal [14–21].

In this paper, we review and focus Malaysia's capacity for using soil mineral carbonation (inorganic  $CO_2$  sequestration) as a tool to mitigate the country's high annual  $CO_2$  emissions, and speculate on how far the technique could contribute to the country's commitment to the COP21 Agreement.

## 2. Malaysia's Position in CO<sub>2</sub> Emissions and CCS

According to a report submitted to the United Nations Framework Convention on Climate Change (UNFCCC) [22], major sources of CO<sub>2</sub> in Malaysia are from:

i. Energy industries: Coal and electrical power industries are the major sources of GHGs emission in Malaysia as well as rest of the world. The total coal consumption for electricity generation in Malaysia is projected to increase from 12.4 million tonnes in 2005 to 36 million tonnes in 2020 [9]. Electricity generation, which contributes 43.40% of total emissions, was discovered to be the largest emitting sector among all sectors.

ii. Transportation sector: This is ranked the second largest GHG emitter among ASEAN countries [23]. This is due to the expansion of conurbation areas such as Kuala Lumpur, where the population is estimated to reach 10 million by 2020 [24]. As a result, motorisation in Malaysia increased five-fold over the past three decades, and proliferation of automobiles is a key contributor towards emission of GHGs [25].

iii. Manufacturing and construction industries: These sectors come in as third for production of GHGs. Malaysia is one of the major manufacturing hubs in ASEAN countries, and remarkable development in this sector is accompanied by high atmospheric CO<sub>2</sub> concentrations [26]. Generally, there are four sources of GHGs emission in manufacturing and construction sectors: (i) Manufacture and transportation of building materials, (ii) Energy consumption of construction equipment, (iii) Energy consumption of processing resources and (iv) Disposal of construction wastes [27]. Figure 1 shows the different sectors for CO<sub>2</sub> emission in Malaysia [22].

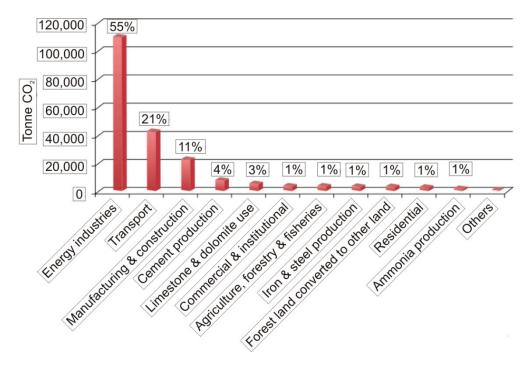


Figure 1. Sectors contributing to CO<sub>2</sub> emissions in Malaysia [22].

According to Malaysia's Intended Nationally Determined Contribution (INDC) [28], by 2030, the country is required to reduce the carbon level to 45% relative to the emissions intensity in 2005. The Malaysian government is persistent in utilising the low carbon industry sector and strengthening the use of green technology through the Government Green Procurement (GGPA) programme. Through the initiatives, the Green Technology Financing Scheme (GTFS) was introduced in 2009, with RM1.5 billion funding to support industries in Malaysia to adopt green technologies. Following the implementation of GTFS scheme in Malaysia, greenhouse gas emission results in 2013 showed a reduction of 94,810 tonnes CO<sub>2</sub> through the applications of green technology. Following the success of the first phase, the second phase was introduced, with allocation of RM5 billion to support industries between 2018 and 2022. Since the power generation sector is the main contributor to the country's overall CO<sub>2</sub> emissions, the Malaysian government's focus is to mitigate excessive emissions generated by this sector. To address this, in 2008 the Malaysian government established Efficient Management of Electrical Energy Regulation (EMEER) to reduce electricity consumption in Malaysia [10].

Malaysia is setting up sustainable regulations aligned with the moving forward strategy 'Transformasi Nasional 2050', known as TN50. The main objective of TN50 is to ensure achievements of the country's Vision 2020 goals, which is to place Malaysia as the top 20 biggest economies in the world, as a model state for management issues related to climate change, and provide affordable and

Geosciences **2018**, *8*, 483 4 of 14

clean energy by 2050. In addition, Malaysia, as an ASEAN country, has adopted the ASEAN Plan of Action for Energy Cooperation (APAEC), and is targeting to reduce the ASEAN energy intensity by 20% in 2020 and 30% by 2050 [10].

Efforts have been made to implement CCS techniques in Malaysia, in line with the TN50 objectives to mitigate impacts of country's  $CO_2$  emissions. The majority of the works towards CCS have been limited to small-scale research projects, however, two existing commercial-scale CCS-related projects in Malaysia are summarised below:

- i. K5 Strategic Technology Project: Malaysia is known as the second-biggest oil producer in Southeast Asia, and the country's national oil company, Petronas, is taking part in CCS development to revive the K5 sour gas project in shallow waters off Sarawak through deployment of carbon capture technologies. The K5 project began in 1970, and contains a gas reservoir of approximately 21 trillion cubic feet. The K5 project gas processing is associated with high CO<sub>2</sub> emissions and, therefore, Petronas has introduced 'K5 Strategic Technology Project' as a pilot scheme to tackle the issues associated with the reservoir's CO<sub>2</sub> emission using CCS technologies. The company is aiming to manufacture and install the first-ever specially built CO<sub>2</sub> processing platform in Malaysia by 2022. The platform will have a hull weighing 11,000 tonnes, and the upper part of offshore, topsides of 9000 tonnes with the attached facilities, are designed to capture CO<sub>2</sub> and transport it into the same offshore reservoir below the seabed [29]. The topside of the platform literally consists of the oil production plant, accommodation board, and drilling rig [29].
- ii. TNB Janamanjung Project: One of the initiatives by Malaysia is application of CCS in coal-fired power stations at TNB Janamanjung, built on a man-made island located in Seri Manjung, Perak. By using post-combustion CCS technology, approximately 85%–95% (8.5–9.5 million tonne CO<sub>2</sub> year<sup>-1</sup>) of the CO<sub>2</sub> is captured and compressed from the processed plant. Later, the compressed CO<sub>2</sub> is transported using an alternative line along the PETRONAS Peninsular Gas Utilization (PGU) project to transfer the captured CO<sub>2</sub> offshore in Terengganu [30]. It is estimated that the PGU system extends to over 1700 km, and the compressed CO<sub>2</sub> can be transferred to the west coast of Peninsular Malaysia where oil and gas exploration is being conducted for geological storage, especially for enhanced oil recovery (EOR).

There is no doubt that CCS technology provides abundant opportunities for CO<sub>2</sub> removal. However, several challenges need to be identified for continuous use of the technique in large and commercial scale. The first and most important challenge is the cost associated with CCS. It is estimated that the overall costs for CCS are within US\$30–70 per tonne CO<sub>2</sub>, with the separation and compression processes accounting for over 75% of overall CCS project costs, whereby underground storage accounts for US\$3–10 per tonne CO<sub>2</sub> [31,32]. Accordingly, to implement the CCS technique at a national level, a significant financial contribution needs to be secured by the Malaysian government. Provision of transportation facilities and infrastructures required for transportation of CO<sub>2</sub> to storage sites are hurdles that have an impact on effectiveness of this technique. The choices of transportation depend on the distance between field and storage sites, available infrastructure, available cost, and geography or geology of the route. Therefore, technical aspects must be considered when contemplating the CCS projects in Malaysia. In terms of storage sites, reservoirs need to be safe from leaking. The safety precautions must be identified properly, and the projects must be handled by an expert to ensure the successful deployment of the CCS projects.

The CCS projects in Malaysia are still at an embryonic stage. In an effort to raise awareness, CCS courses have been introduced to Malaysian universities in 2013, and are currently being offered at Universiti Tenaga Nasional (UNITEN), Institute of Technology Petronas (UTP), National University of Malaysia (UKM), and the University of Technology Malaysia (UTM) [33].

Geosciences 2018, 8, 483 5 of 14

## 3. Pedogenic CO<sub>2</sub> Sequestration

The atmosphere contains approximately 730 G tonnes of carbon. By contrast, the oceans provide a reservoir for approximately 38,000 G tonnes [34]. The upper 100 cm of soil is reservoir for 1500 G tonnes of organic carbon and land-plant biomass (belowground within root system) contains  $\sim$ 500 G tonnes carbon [34]. In addition, soil's pedogenic carbonate is a reservoir for 750–950 G tonnes of carbon [35]. Accordingly, soil organic matter and pedogenic carbonates, as the hosts for organic and inorganic carbon, respectively, are a major terrestrial sink for carbon, containing three times as much carbon as aboveground plant biomass [15]. Engineering soil, through enhancement of the soil organic carbon and pedogenic carbonate stocks, has proven to be a potentially effective method for atmospheric  $CO_2$  removal [14–21,36–42]. Sequestration of  $CO_2$ , in soil, occurs in organic and inorganic formats.

# 3.1. Organic CO<sub>2</sub> Sequestration

Soil organic carbon stock and its potential for further sequestration has been well researched, and the focus for achieving goals which were set to tackle issues associated with global warming outlined in frameworks, such as COP21 [43–50]. Sequestration of carbon organically, in addition to the conservation of existing soil carbon stocks, are two important pathways contributing significantly towards the COP21's target of maintaining global temperature increase less than 2 °C [50]. Forest soils are an important reservoir for organic carbon sequestration in Malaysia. Malaysia has a forest area of 17.7 M ha, of a total land area 330,803 km² (33 M ha), i.e., 53.64% is forest compared to the total area of Malaysia. It is possible for Malaysia to sequester 15 tonnes of  $CO_2$  ha $^{-1}$ year $^{-1}$  in forest soil [7]. However, climate change and stagnating crops may lead to reduction in soil organic carbon stock [51], making it an uncertain sink for carbon and an unstable method for atmospheric  $CO_2$  removal.

# 3.2. Inorganic CO<sub>2</sub> Sequestration

Soil inorganic  $CO_2$  sequestration, as inorganic carbon, has the potential to be an effective method for atmospheric CO<sub>2</sub> removal [14,15,18-20,36,37]. Calcium content and availability of CO<sub>2</sub> in the substrate are two important factors controlling the formation of pedogenic carbonates which are predominately composed of the mineral calcite (CaCO<sub>3</sub>). Calcium is derived naturally from the weathering of silicate minerals (plagioclase feldspars, pyroxenes etc.) that commonly occur in basic igneous rocks (e.g., basalts) or artificial calcium silicate and hydroxide minerals present in concrete and cement [14,16,52]. In addition, fly and bottom ashes contain calcium (as CaO) required for inorganic CO<sub>2</sub> sequestration [53]. Decomposition of organic acid anions [18], which combine with other biogenic carbon inputs in soils to produce CO<sub>2</sub> as the ultimate product of aerobic decomposition [36]. A proportion of the CO<sub>2</sub>, depending on the pH, partitions into the soil solution as bicarbonate or carbonate [36]. Calcium in solution (derived from weathering of silicate minerals) reacts with dissolved CO<sub>2</sub> to form carbonates (CaCO<sub>3</sub>). This leads to removal of atmospheric CO<sub>2</sub> and formation of CaCO<sub>3</sub> in soil as a stable sink [14,16–20,36,54]. Power et al. [55] identified ultramafic and mafic mine tailings as alternative materials required for inorganic CO<sub>2</sub> sequestration. Unlike organic CO<sub>2</sub> sequestration, once CaCO<sub>3</sub> is formed in soil, it stays as a stable sink which could only be removed naturally through dissolution, and entering surface and groundwater systems [36].

Inorganic carbon sequestration in Malaysia has not been extensively studied, and the country's potential for using inorganic carbon sequestration for atmospheric CO<sub>2</sub> removal is unknown. Only recently, Syed Hasan et al. [56] investigated the potential of gold mine waste for inorganic carbon sequestration, and realised that there is great potential for the materials to be used as a tool for passive CO<sub>2</sub> removal. Malaysia is a developing country, and will see a large volume of demolition and construction activities, where the impact and opportunity for a national carbon budget in redevelopment can be considered. Demolition provides calcium-rich material, and redevelopment provides a unique chance to integrate inorganic CO<sub>2</sub> sequestration into the design of future structures. In addition to the demolition waste in Malaysia, other sources of Ca include by-products such as coal

ash (bottom and fly ash) waste from power plants, and basalt quarry fine containing calcium-rich material necessary for inorganic  $CO_2$  sequestration.

Sufficient soluble  $Ca^{2+}/Mg^{2+}$  in soil results in carbonate precipitation [57]. Accordingly, one of the main limiting factors for inorganic  $CO_2$  sequestration is exhaustion of calcium sources in soil. According to Jorat et al. [54], occupation of soil void spaces as a result of calcite precipitation would eventually lead to termination of inorganic  $CO_2$  sequestration process. Concerns have been raised on flood risk at urban sites engineered for inorganic  $CO_2$  sequestration, due to soil pore spaces clogging as a result of calcite precipitation [36]. Reduction of permeability was observed on soil samples treated artificially with microbial induced calcite precipitation (MICP) (e.g., Al Qabany and Soga [58]). Due to annual heavy rainfalls in Malaysia, which could lead to flooding, prior to large-scale deployment of the inorganic  $CO_2$  sequestration technique, flooding risk assessment must be conducted using pilot studies. MICP has also been demonstrated to increase ground strength through cementation of soil particles [58]. Where enhancement of ground strength is preferable, inorganic  $CO_2$  sequestration could be designed into the ground to couple  $CO_2$  capture with ground improvement. For engineering practice, inorganic  $CO_2$  sequestration could be used as a natural process to stabilise strength in soil.

Groundwater stores carbonate in the form of bicarbonate [59], which is considered a long-term effect of inorganic CO<sub>2</sub> sequestration. Formation of CaCO<sub>3</sub> in soil, as a result of inorganic CO<sub>2</sub> sequestration, is known to increase the environment's pH [15–20], and can be used as a technique to reduce acidity of soils with low pH. If inorganic CO<sub>2</sub> sequestration is implemented at a large scale, monitoring of groundwater bicarbonate is required. Contamination is an important factor which should be taken into account when choosing source of calcium required for inorganic CO<sub>2</sub> sequestration. This is particularly important when dealing with demolition concrete waste, and fly and bottom ashes, as contaminated leachate might cause extensive (and often permanent) groundwater contamination. Contamination analysis must be conducted on the calcium sources prior to deployment.

# 4. Industrial Waste in Malaysia

Various types of industrial wastes are being generated in Malaysia, and only three types related to this study are being presented in this section. Chemical composition of selected samples from the industrial wastes are presented in Table 1.

**Table 1.** Typical chemical composition of demolition concrete waste, basalt, and fly and bottom ash samples from Malaysia.

Element	Demolition Concrete Waste (wt %) [60]	Basalt (wt %) [61]	Fly Ash (wt %) [62]	Bottom Ash (wt %) [62]
CaO	$70.88 \pm 9.22$	11.08	4.81	9.8
$SiO_2$	$20.68 \pm 6.47$	47.17	51.8	42.7
$Al_2O_3$	$3.43 \pm 1.52$	16.78	26.5	23
FeO	-	8.89	-	-
$Fe_2O_3$	$1.38 \pm 0.73$	-	8.5	17
Na <sub>2</sub> O	$0.06 \pm 0.01$	2.2	0.67	0.29
MgO	$1.99 \pm 0.19$	8.07	1.1	1.54
K <sub>2</sub> O	$0.67 \pm 0.13$	1.26	3.27	0.96
$TiO_2$	$0.11 \pm 0.04$	1.13	1.38	1.64
MnO	$0.06 \pm 0.55$	0.11	-	
$P_2O_5$	$0.06 \pm 0.02$	0.1	0.9	1.04
Rb <sub>2</sub> O	$0.01 \pm 0$	-	-	-
SrO	$0.04 \pm 0.01$	-	-	-
$ZrO_2$	$0.02 \pm 0.01$	=	-	-
BaO	-		0.12	0.19
$SO_3$	$0.61 \pm 0.47$	-	0.6	1.22
LOI	-	3.02	-	-

Geosciences 2018, 8, 483 7 of 14

#### 4.1. Demolition Concrete Waste

Development in the construction sector has resulted in the significant production of solid waste which, in turn, could create environmental-related issues. Due to rapid development in the economy, population growth, and inadequate disposal land and infrastructure, solid waste has become Malaysia's most critical environmental issue [63]. In particular, inappropriate deposition of demolition waste has resulted in serious environmental issues in Malaysia [64]. Only 15% of overall demolition and construction waste in Malaysia are recycled annually, which is far behind countries such as Singapore, Germany, and South Korea, with the rate of 50%–75% [65]. Malaysia generates 26,000 tonnes of construction and demolition waste daily, leading to the generation of nearly 10 million tonnes of wastes annually [65]. According to Nagapan et al. [66], a study of 30 demolition sites in Malaysia showed that concrete comprised 12.3% of the total demolition waste. Considering the total annual generation of demolition waste, a minimum of 1.2 million tonnes of concrete waste is being produced in Malaysia annually. Demolition concrete waste comes in various sizes and shapes, depending on its origin. The size varies from clay size to large blocks which, in the latter case, must be further crushed to provide larger surface area.

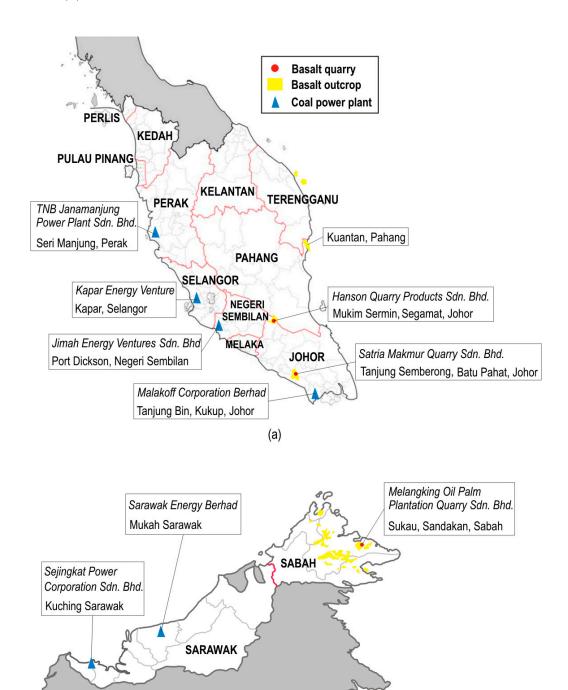
#### 4.2. Basalt Quarry Fine

Basalt quarries for aggregate production generate two types of residues: (i) aggregates produced during crushing and milling operations, which are used as fine aggregates by the construction sector, and (ii) fines remain from crushing and milling operations, which have no commercial value to be used as aggregate for construction. Based on the information acquired from operational basalt quarries in Malaysia, clay size fines account for 18–20 wt % of the aggregate production, and are commonly being deposited in large quantities at quarries. Recommendations have been given on suitability of basalt quarry fines for Portland cement production (e.g., [67]); however, where limestone is abundant (as in Malaysia), basalt quarry fines often would not be used as raw material for Portland cement production.

In the Malay Peninsula, basalt is found as occasional outcrops [68], for example, at Baserah, Pahang in Malaysia [69] (Figure 2). According to Hamdan et al. [70], basalt in Pahang spreads over an area of around 30,000 ha. Ghani et al. [61] reported existence of basalt in Kuantan, Perhentian, Redang Islands, and mainland Terengganu (Figure 2). According to the [71], three active basalt quarries operate in Malaysia (two in Johor and one in Sabah) (Figure 2). Basalt aggregate production from these increased significantly from 27,400 tonnes in 2010, to 344,930 tonnes in 2017 [71]. Considering the 18–20 wt % quarry fine production, it is estimated that Malaysia produces 62,000–69,000 tonnes of basalt quarry fines, with no current commercial value.

# 4.3. Coal Ash

Production of electricity in Malaysia highly relies on six coal-fired thermal power plants [72] (Figure 2), which are associated with the generation of fly and bottom ashes in significant volumes. In Malaysia, fly and bottom ashes are commonly deposited in uncovered landfills, which impose significant environmental issues. According to Rafieizonooz et al. [73], Malaysia produces 6.8 and 1.7 million tonnes of fly and bottom ashes annually, respectively. Studies show the suitability of fly and bottom ashes, in Malaysia, for construction purposes (e.g., [74–77]). However, these are still at research level, and fly and bottom ashes have not been utilised at industrial scale in construction activities in Malaysia. Muhardi et al. [62] reported fly ash size <0.04 mm and bottom ash size 0.07< <11 mm.



**Figure 2.** Map of (a) Peninsular Malaysia (adopted from https://openclipart.org/detail/290360/peninsular-malaysia-blank-map) and (b) East Malaysia showing locations for basalt quarry, basalt outcrop, and coal power plant. Information regarding basalt quarries and outcrops was derived from Department of Mineral and Geoscience Malaysia [71] and, regarding coal power plants, was derived from https://endcoal.org/global-coal-plant-tracker/.

(b)

# 5. Malaysia's Capacity for Using Soil Mineral Carbonation

Malaysia, as a developing country, is anticipating an increase in construction and demolition activities in the future [78]. In Europe, the life of concrete buildings is 40–50 years, in contrast to 25 years in a tropical country such as Malaysia. Considering Malaysia's increasing  $CO_2$  emission and

growing construction and demolition waste figures, there is an abundant scope for inorganic CO<sub>2</sub> sequestration to be exploited and implemented, in order to utilise demolition waste for atmospheric CO<sub>2</sub> reduction. Washbourne et al. [19] investigated a site in the United Kingdom containing demolition waste, and showed that the top 10 cm of soil at the site sequestered 85 tonnes CO<sub>2</sub> ha<sup>-1</sup>year<sup>-1</sup>. Manning et al. [79] measured sequestration capacity of 6 tonnes  $CO_2$  ha<sup>-1</sup>year<sup>-1</sup> for the top 10 cm of soil at the site containing 50% dolerite fines (which has similar material composition to basalt) and 50% compost. Considering annual generation of nearly 1.2 million tonnes of demolition concrete wastes [65] and 62,000–69,000 tonnes of basalt quarry fines [80], these materials can be used for inorganic CO<sub>2</sub> sequestration in Malaysia, if the function is designed into the first 10 cm of land associated with the country's future developments. Taking the availability of demolition concrete waste and basalt quarry fines into consideration, and assuming an application rate of 100% demolition concrete waste and 50% basalt + 50% compost, Malaysia has the capacity to annually establish 1043 and 85 ha of land suitable for inorganic CO<sub>2</sub> sequestration using demolition concrete waste and basalt quarry fine, respectively. Accordingly, taking into account the sequestration rates for demolition concrete waste and basalt quarry fine, 89,000 tonnes of CO<sub>2</sub> can be captured annually in Malaysia (Table 2). This figure accounts for 0.03% of the country's projected total CO<sub>2</sub> emission in 2020.

**Table 2.** Annual production for the three types of industrial waste in Malaysia suitable for inorganic CO<sub>2</sub> sequestration, area of land that could be engineered, and the country's annual CO<sub>2</sub> sequestration capacity.

Material	Annual Production (tonne)	Engineered Area (ha)	Annual CO <sub>2</sub> Sequestration Capacity (tonne)
Demolition concrete waste	1,200,000	1043	88,655
Basalt quarry fine	62,000-69,000	85	510
Fly ash	6,800,000	7907	176,800
Bottom ash	1,700,000	1143	331,500
		Total = 10,178	Total = 597,465

Annual CO<sub>2</sub> sequestration capacity of *in-situ* soils containing fly ash and bottom ash have not been investigated. However, based on a laboratory study, Montes-Hernandez et al. [53] reported maximum CO<sub>2</sub> sequestration capacity of 26 kg for one tonne of fly ash. In addition, based on another laboratory study, Kim and Lee [81] observed maximum CO<sub>2</sub> sequestration capacity of 195 kg for one tonne of bottom ash. If the 6.8 and 1.7 million tonnes of fly and bottom ashes, respectively, being produced annually in Malaysia [73] would be used for the first 10 cm of land associated with the country's future developments (e.g., using lands associated with construction of highway [14,82]), more than 9000 ha of land could be designed for inorganic CO<sub>2</sub> sequestration, and 500,000 tonnes of CO<sub>2</sub> could be captured. This figure accounts for 0.2% of the country's projected total CO<sub>2</sub> emission in 2020.

Temperature is a parameter that controls the kinetics of mineral carbonation reaction. An increase in temperature is known to increase reaction kinetics [83]. Accordingly, in a country like Malaysia, with a tropical climate, the mineral carbonation rate should be higher than compared to the United Kingdom. To accurately quantify the differences in rate, similar small-scale experiments to those in the United Kingdom should be conducted, and annual sequestration capacity of the by-products should be measured.

#### 6. Conclusions

After signing the COP21, Malaysia is looking for solutions to tackle developing annual  $CO_2$  emissions. Carbon capture and storage is very new to the country, and apart from two notable examples, remaining efforts in the direction of  $CO_2$  sequestration have been limited to small-scale research studies. In this study, our hypothetical analysis shows that inorganic  $CO_2$  sequestration could be a suitable solution for Malaysia's  $CO_2$  emissions. We have identified three main by-products in Malaysia containing calcium, which is required for inorganic  $CO_2$  sequestration, namely, demolition concrete waste, basalt quarry fine, and fly and bottom ashes. Using the by-products for climate mitigation also reduces the country's annual waste productions and minimises negative environmental impacts associated with often unsuitable deposition of the by-products. Our analysis shows that 10,178 ha of readily available lands, associated with the country's future developments, could be engineered annually using the by-products being produced in Malaysia, leading to the sequestration of 597,465 tonnes of  $CO_2$  year<sup>-1</sup>. This accounts for 0.23% of the country's projected total  $CO_2$  emission in 2020, with a potential annual economic benefit of  $\{4,700,000\}$ .

**Author Contributions:** Conceptualisation, M.E.J. and A.M. and D.A.C.M.; formal analysis, M.E.J.; investigation, M.E.J. and M.A.A. and N.Z. and S.N.J.; resources, M.A.A. and N.Z. and S.N.J.; data curation, M.E.J.; writing—original draft preparation, M.E.J. and M.A.A and N.Z. and S.N.J.; writing—review and editing, A.M. and D.A.C.M; visualisation, M.E.J.; supervision, M.E.J. and A.M.

**Funding:** Scottish Alliance for Geosciences, Environment and Society (SAGES) covered travel expenses for M.E.J. to visit UTM in Malaysia which initiated this research collaboration between the UK and Malaysia.

**Acknowledgments:** We would like to thank support from Department of Mineral and Geosciences Malaysia during preparation of this paper. We appreciate financial support from the Scottish Alliance for Geosciences, Environment and Society (SAGES). Mohammed A.M. Al-Bared reviewed the text and references.

Conflicts of Interest: The authors declare no conflict of interest.

## References

- International Energy Agency. Decoupling of Global Emissions and Economic Growth Confirmed. 2016. Available online: https://www.iea.org/newsroom/news/2016/march/decoupling-of-global-emissions-and-economic-growth-confirmed.html (accessed on 13 November 2018).
- 2. Stern, N. The Economics of Climate Change; Oxford University Press: Oxford, UK, 2006.
- 3. The Institute of Energy Economics, Japan. IEEJ Outlook 2018. Available online: https://eneken.ieej.or.jp/data/7748.pdf (accessed on 11 September 2018).
- 4. U.S. Energy Information Administration. *International Energy Outlook 2011: With Projections to 2035;* U.S. Energy Information Administration: Washington, DC, USA, 2011.
- Leeson, D.; Mac Dowell, N.; Shah, N.; Petit, C.; Fennell, P. A Techno-Economic Analysis and Systematic Review of Carbon Capture and Storage (CCS) Applied to the Iron and Steel, Cement, Oil Refining and Pulp and Paper Industries, as Well as Other High Purity Sources. *Int. J. Greenh. Gas Control* 2017, 61, 71–84.
  [CrossRef]
- 6. Steve, A.L. *Implementation of MS1525 & Low Carbon Buildings Strategic & Affordable Way to Reduce CO*<sub>2</sub> *Emissions for Building Sector*; Sustainable Energy Development Authority (SEDA): Putrajaya, Malaysia, 2017.
- 7. Safaai, N.S.M.; Noor, Z.Z.; Hashim, H.; Ujang, Z.; Talib, J. Projection of CO<sub>2</sub> Emissions in Malaysia. *Environ. Prog. Sustain. Energy* **2011**, *30*, 658–665. [CrossRef]
- 8. Salahudin, S.N.; Abdullah, M.M.; Newaz, N.A. Emissions: Sources, Policies and Development in Malaysia. *Int. J. Educ. Res.* **2013**, *1*, 1–12.
- 9. Othman, M.; Zakaria, R.; Fernando, W. Strategic Planning on Carbon Capture from Coal Fired Plants in Malaysia and Indonesia: A Review. *Energy Policy* **2009**, *37*, 1718–1735. [CrossRef]
- Energy Commission. Energy Malaysia: Towards a World-Class Energy Sector 2017. Available online: https://www.st.gov.my/ms/contents/publications/energyMalaysia/EM12%20Nov%202017%20v2.pdf (accessed on 13 November 2018).

11. Paul, F.; Anthony, A.O.C.; Hisashi, I.; William, M.; Jose, M. Chapter 1: Introduction. In IPCC Special Report on Carbon Dioxide Capture and Storage 2011. Available online: https://www.ipcc.ch/report/carbon-dioxide-capture-and-storage/ (accessed on 13 November 2018).

- 12. Nor, N.H.M.; Selamat, S.N.; Rashid, M.H.A.; Ahmad, M.F.; Jamian, S.; Kiong, S.C.; Yokoyama, S. Carbon Sequestration and Carbon Capture and Storage (CCS) in Southeast Asia. *J. Phys. Conf. Ser.* **2016**, 725, 012010.
- 13. Intergovernmental Panel on Climate Change (IPCC). Summary for Policymakers. In Global Warming of 1.5 °C. Available online: https://www.ipcc.ch/sr15/chapter/summary-for-policy-makers/ (accessed on 13 November 2018).
- 14. Jorat, M.E.; Kolosz, B.W.; Goddard, M.A.; Sohi, S.P.; Akgun, N.; Dissanayake, D.; Manning, D.A. Geotechnical Requirements for Capturing CO<sub>2</sub> through Highways Land. *Int. J. GEOMATE* **2017**, *13*, 22–27. [CrossRef]
- 15. Manning, D. Biological Enhancement of Soil Carbonate Precipitation: Passive Removal of Atmospheric CO<sub>2</sub>. *Mineral. Mag.* **2008**, 72. [CrossRef]
- 16. Manning, D.A.; Renforth, P. Passive Sequestration of Atmospheric CO<sub>2</sub> through Coupled Plant-Mineral Reactions in Urban Soils. *Environ. Sci. Technol.* **2012**, *47*, 135–141. [CrossRef]
- 17. Renforth, P.; Leake, J.R.; Edmondson, J.; Manning, D.A.; Gaston, K.J. Designing a Carbon Capture Function into Urban Soils. *Proc. ICE-Urban Des. Plan.* **2011**, *164*, 121–128. [CrossRef]
- 18. Renforth, P.; Manning, D.; Lopez-Capel, E. Carbonate Precipitation in Artificial Soils as a Sink for Atmospheric Carbon Dioxide. *Appl. Geochem.* **2009**, *24*, 1757–1764. [CrossRef]
- 19. Washbourne, C.-L.; Lopez-Capel, E.; Renforth, P.; Ascough, P.L.; Manning, D.A. Rapid Removal of Atmospheric CO<sub>2</sub> by Urban Soils. *Environ. Sci. Technol.* **2015**, *49*, 5434–5440. [CrossRef] [PubMed]
- 20. Washbourne, C.-L.; Renforth, P.; Manning, D. Investigating Carbonate Formation in Urban Soils as a Method for Capture and Storage of Atmospheric Carbon. *Sci. Total Environ.* **2012**, *431*, 166–175. [CrossRef] [PubMed]
- 21. Kolosz, B.; Goddard, M.; Jorat, M.E.; Aumonier, J.; Sohi, S.; Manning, D.A.C. A Sustainability Framework for Engineering Carbon Capture Soil in Transport Infrastructure. *Int. J. Transp. Dev. Integr.* **2017**, *1*, 74–83. [CrossRef]
- 22. Ministry of Natural Resources and Environment Malaysia. Malaysian Biennial Update Report to the UNFCCC. 2015. Available online: https://unfccc.int/files/national\_reports/non-annex\_i\_parties/biennial\_update\_reports/application/pdf/malbur1.pdf (accessed on 13 November 2018).
- 23. Ghadimzadeh, A.; Makmom, A.A.; Hosea, M.K.; Asgari, N.; Shamsipour, R.; Askari, A.; Narany, T.S. Review on CO<sub>2</sub> Emission from Transportation Sector in Malaysia. *IOSR J. Environ. Sci. Toxicol. Food Technol.* **2015**, *9*, 61–70.
- 24. Kwan, S.C.; Tainio, M.; Woodcock, J.; Sutan, R.; Hashim, J.H. The Carbon Savings and Health Co-Benefits from the Introduction of Mass Rapid Transit System in Greater Kuala Lumpur, Malaysia. *J. Transp. Health* **2017**, *6*, 187–200. [CrossRef]
- 25. Shahid, S.; Minhans, A.; Puan, O.C. Assessment of Greenhouse Gas Emission Reduction Measures in Transportation Sector of Malaysia. *J. Teknol.* **2014**, *70*, 1–8. [CrossRef]
- 26. Chin, M.-Y.; Puah, C.-H.; Teo, C.-L.; Joseph, J. The Determinants of CO<sub>2</sub> Emissions in Malaysia: A New Aspect. *Int. J. Energy Econ. Policy* **2018**, *8*, 190–194.
- 27. Klufallah, M.M.; Nuruddin, M.F.; Khamidi, M.F.; Jamaludin, N. Assessment of Carbon Emission Reduction for Buildings Projects in Malaysia—A Comparative Analysis. *E3S Web Conf.* **2014**, *3*, 01016. [CrossRef]
- 28. Government of Malaysia. Intended Nationally Determined Contribution of the Government of Malaysia. Available online: https://www4.unfccc.int/sites/submissions/INDC/Published%20Documents/Malaysia/1/INDC%20Malaysia%20Final%2027%20November%202015%20Revised%20Final%20UNFCCC.pdf (accessed on 13 November 2018).
- 29. Russel, S.W. Petronas to Revive K5 as Part of CCS Development. 2018. Available online: https://www.upstreamonline.com/hardcopy/1459531/petronas-to-revive-k5-as-part-of-ccs-development (accessed on 13 November 2018).
- 30. Oh, T.H. Carbon Capture and Storage Potential in Coal-Fired Plant in Malaysia—A Review. *Renew. Sustain. Energy Rev.* **2010**, *14*, 2697–2709. [CrossRef]
- 31. Rubin, E.S.; Rao, A.B. Uncertainties in CO<sub>2</sub> Capture and Sequestration Costs. In *Greenhouse Gas Control Technologies*—6th International Conference; Elsevier: Amsterdam, The Netherlands, 2003.
- 32. Simbeck, D. The 10-50 Solution: Technologies and Policies for a Low-Carbon Future. *Energy Policy* **2004**, *57*, 2266–2278.

33. Gibson, A. Successful Introduction of Ccs 101 Course into Malaysian Universities. 2016. Available online: https://www.globalccsinstitute.com/insights/authors/AliceGibson/2016/01/14/successful-introduction-ccs-101-course-malaysian-universities (accessed on 18 October 2018).

- 34. Smith, P. Carbon Sequestration in Croplands: The Potential in Europe and the Global Context. *Eur. J. Agron.* **2004**, *20*, 229–236. [CrossRef]
- 35. Schlesinger, W.H. The Formation of Caliche in Soils of the Mojave Desert, California. *Geochim. Cosmochim. Acta* **1985**, 49, 57–66. [CrossRef]
- 36. Jorat, M.; Goddard, M.; Kolosz, B.; Sohi, S.; Manning, D. Sustainable Urban Carbon Capture: Engineering Soils for Climate Change (Success). In Proceedings of the 16th European Conference on Soil Mechanics and Geotechnical Engineering (XVI ECSMGE), Edinburgh, UK, 13–17 September 2015.
- 37. Landi, A.; Mermut, A.; Anderson, D. Origin and Rate of Pedogenic Carbonate Accumulation in Saskatchewan Soils, Canada. *Geoderma* **2003**, *117*, 143–156. [CrossRef]
- 38. Mbow, C.; Smith, P.; Skole, D.; Duguma, L.; Bustamante, M. Achieving Mitigation and Adaptation to Climate Change through Sustainable Agroforestry Practices in Africa. *Curr. Opin. Environ. Sustain.* **2014**, *6*, 8–14. [CrossRef]
- 39. Powlson, D.S.; Whitmore, A.P.; Goulding, K.W. Soil Carbon Sequestration to Mitigate Climate Change: A Critical Re-Examination to Identify the True and the False. *Eur. J. Soil Sci.* **2011**, *62*, 42–55. [CrossRef]
- 40. Schmidt, M.W.; Torn, M.S.; Abiven, S.; Dittmar, T.; Guggenberger, G.; Janssens, I.A.; Manning, D.A. Persistence of Soil Organic Matter as an Ecosystem Property. *Nature* **2011**, 478, 49. [CrossRef]
- 41. Stockmann, U.; Adams, M.A.; Crawford, J.W.; Field, D.J.; Henakaarchchi, N.; Jenkins, M.; Singh, K. The Knowns, Known Unknowns and Unknowns of Sequestration of Soil Organic Carbon. *Agric. Ecosyst. Environ.* **2013**, *164*, 80–99. [CrossRef]
- 42. Wutzler, T.; Reichstein, M. Soils Apart from Equilibrium? *Consequences for Soil Carbon Balance Modelling. Biogeosci. Discuss.* **2006**, *3*, 1679–1714.
- 43. Crowther, T.W.; Todd-Brown, K.E.; Rowe, C.W.; Wieder, W.R.; Carey, J.C.; Machmuller, M.B.; Allison, S.D. Quantifying Global Soil Carbon Losses in Response to Warming. *Nature* **2016**, *540*, 104–108. [CrossRef]
- 44. He, Y.; Trumbore, S.E.; Torn, M.S.; Harden, J.W.; Vaughn, L.J.; Allison, S.D.; Randerson, J.T. Radiocarbon Constraints Imply Reduced Carbon Uptake by Soils During the 21st Century. *Science* **2016**, *353*, 1419–1424. [CrossRef]
- 45. Minasny, B.; Malone, B.P.; McBratney, A.B.; Angers, D.A.; Arrouays, D.; Chambers, A.; Das, B.S. Soil Carbon 4 Per Mille. *Geoderma* **2017**, 292, 59–86. [CrossRef]
- 46. Paustian, K.; Lehmann, J.; Ogle, S.; Reay, D.; Robertson, G.P.; Smith, P. Climate-Smart Soils. *Nature* **2016**, 532, 49. [CrossRef] [PubMed]
- 47. Pries, C.E.H.; Castanha, C.; Porras, R.; Torn, M. The Whole-Soil Carbon Flux in Response to Warming. *Science* **2017**, 355, 1420–1423. [CrossRef] [PubMed]
- 48. Van Groenigen, J.W.; Van Kessel, C.; Hungate, B.A.; Oenema, O.; Powlson, D.S.; Van Groenigen, K.J. Sequestering Soil Organic Carbon: A Nitrogen Dilemma. *Environ. Sci. Technol.* **2017**, *51*, 4738–4739. [CrossRef] [PubMed]
- 49. Zhao, X.; Deng, H.; Wang, W.; Han, F.; Li, C.; Zhang, H.; Dai, Z. Impact of Naturally Leaking Carbon Dioxide on Soil Properties and Ecosystems in the Qinghai-Tibet Plateau. *Sci. Rep.* **2017**, *7*, 3001. [CrossRef] [PubMed]
- 50. Zomer, R.J.; Bossio, D.A.; Sommer, R.; Verchot, L.V. Global Sequestration Potential of Increased Organic Carbon in Cropland Soils. *Sci. Rep.* **2017**, *7*, 15554. [CrossRef] [PubMed]
- 51. Wiesmeier, M.; Poeplau, C.; Sierra, C.A.; Maier, H.; Frühauf, C.; Hübner, R.; Hangen, E. Projected Loss of Soil Organic Carbon in Temperate Agricultural Soils in the 21 St Century: Effects of Climate Change and Carbon Input Trends. *Sci. Rep.* **2016**, *6*, 32525. [CrossRef]
- 52. Kolosz, B.; Goddard, M.; Jorat, M.E.; Sohi, S.; Manning, D.A.C. Developing Lifecycle Inventory Indices for Estimating the Carbon Sequestration of Artificially Engineered Soils and Plants. In Proceedings of the 5th Asian Conference on Sustainability, Energy, and the Environment, Kobe, Japan, 11–14 June 2015.
- 53. Montes-Hernandez, G.; Pérez-López, R.; Renard, F.; Nieto, J.M.; Charlet, L. Mineral sequestration of CO<sub>2</sub> by aqueous carbonation of coal combustion fly-ash. *J. Hazard. Mater.* **2008**, *161*, 1347–1354. [CrossRef]
- 54. Jorat, M.; Kolosz, B.; Sohi, S.; Lopez-Capel, E.; Manning, D.A. Changes in Geotechnical Properties of Urban Soils During Carbonation. In Proceedings of the 15th Pan-American Conference on Soil Mechanics and Geotechnical Engineering, Buenos Ares, Argentina, 15–18 November 2015.

55. Power, I.M.; McCutcheon, J.; Harrison, A.L.; Wilson, S.A.; Dipple, G.M.; Kelly, S.; Southam, G. Strategizing Carbon-Neutral Mines: A Case for Pilot Projects. *Minerals* **2014**, *4*, 399–436. [CrossRef]

- 56. Syed Hasan, S.; Mohd Kusin, F.; Jusop, S.; Mohamat Yusuff, F. Potential of Soil, Sludge and Sediment for Mineral Carbonation Process in Selinsing Gold Mine, Malaysia. *Minerals* **2018**, *8*, 257. [CrossRef]
- 57. Guo, Y.; Wang, X.J.; Li, X.L.; Wang, J.P.; Xu, M.G.; Li, D.W. Dynamics of soil organic and inorganic carbon in the cropland of upper Yellow River Delta, China. *Sci. Rep.* **2016**, *6*, 36105. [CrossRef] [PubMed]
- 58. Al Qabany, A.; Soga, K. Effect of chemical treatment used in MICP on engineering properties of cemented soils. *Géotechnique* **2013**, *63*, 331–339. [CrossRef]
- 59. Monger, H.C.; Kraimer, R.A.; Khresat, S.; Cole, D.R.; Wang, X.; Wang, J. Sequestration of inorganic carbon in soil and groundwater. *Geology* **2015**, *43*, 375–378. [CrossRef]
- 60. Nurhanim, A.A.; Norli, I.; Morad, N.; Khalil, H.P.S.A. Leaching Behavior of Construction and Demolition Waste (Concrete and Gypsum). *Iran. J. Energy Environ.* **2016**, *7*, 203–211.
- 61. Ghani, A.A.; Lo, C.-H.; Chung, S.-L. Basaltic dykes of the Eastern Belt of Peninsular Malaysia: The effects of the difference in crustal thickness of Sibumasu and Indochina. *J. Asian Earth Sci.* **2013**, 77, 127–139. [CrossRef]
- 62. Marto, A.; Kassim, K.A.; Makhtar, A.M.; Wei, L.F.; Lim, Y.S. Engineering characteristics of Tanjung Bin coal ash. *Electron. J. Geotech. Eng.* **2010**, *15*, 1117–1129.
- 63. Manaf, L.A.; Abu Samah, M.A.; Mohd Zukki, N.I. Municipal solid waste management in Malaysia: Practices and challenges. *Waste Manag.* **2009**, *29*, 2902–2906. [CrossRef] [PubMed]
- 64. Saadi, N.; Ismail, Z.; Alias, Z. A review of construction waste management and initiatives in Malaysia. *J. Sustain. Sci. Manag.* **2016**, *11*, 101–114.
- 65. Esa, M.R.; Halong, A.; Rigamonti, L. Strategies for Minimizing Construction and Demolition Wastes in Malaysia. *Resour. Conserv. Recycl.* **2017**, 120, 219–229. [CrossRef]
- 66. Nagapan, S.; Rahman, I.A.; Asmi, A. Construction Waste Management: Malaysian Perspective. In Proceeding of the International Conference on Civil and Environmental Engineering Sustainability (IConCEES), Thistle Hotel, Johor, Malaysia, 3–5 April 2012.
- 67. Andrade, F.D.; Pecchio, M.; Bendoraitis, D.; Montanheiro, T.; Kihara, Y. Basalt Mine-Tailings as Raw-Materials for Portland Clinker. *Cerâmica* **2010**, *56*, 39–43. [CrossRef]
- 68. Gobbett, D.G. Geological Map of the Malay Peninsula; Geological Society of Malaysia: Kuala Lumpur, Malaysia, 1972.
- 69. Shamshuddin, J.; Kapok, J. Effect of Ground Basalt on Chemical Properties of an Ultisol and Oxisol in Malaysia. *Pertan. J. Trop. Agric. Sci.* **2010**, *33*, 7–14.
- 70. Hamdan, J.; Ruhana, B.; McRae, S. Characteristics of Regolith Developed on Basalt in Pahang, Malaysia. *Commun. Soil Sci. Plant Anal.* **2000**, *31*, 981–993. [CrossRef]
- 71. Department of Mineral and Geoscience Malaysia. *Industrial Mineral Production Statistics and Directory of Producers in Malaysia*; Department of Mineral and Geoscience Malaysia: Putrajaya, Malaysia, 2017.
- 72. Marto, A.; Tan, C.S. Properties of Coal Bottom Ash from Power Plants in Malaysia and Its Suitability as Geotechnical Engineering Material. *J. Teknol.* **2016**, *78*, 1–10. [CrossRef]
- 73. Rafieizonooz, M.; Mirza, J.; Salim, M.R.; Hussin, M.W.; Khankhaje, E. Investigation of Coal Bottom Ash and Fly Ash in Concrete as Replacement for Sand and Cement. *Constr. Build. Mater.* **2016**, *116*, 15–24. [CrossRef]
- 74. Jorat, M.E.; Marto, A.; Namazi, E.; Amin, M. Engineering Characteristics of Kaolin Mixed with Various Percentages of Bottom Ash. *Electron. J. Geotech. Eng.* **2011**, *16*, 841–850.
- 75. Latifi, N.; Marto, A.; Rashid, A.S.A.; Yii, J.L.J. Strength and Physico-Chemical Characteristics of Fly Ash–Bottom Ash Mixture. *Arabian J. Sci. Eng.* **2015**, *40*, 2447–2455. [CrossRef]
- 76. Marto, A.; Hassan, M.A.; Makhtar, A.M.; Othman, B.A. Shear Strength Improvement of Soft Clay Mixed with Tanjung Bin Coal Ash. *APCBEE Procedia* **2013**, *5*, 116–122. [CrossRef]
- 77. Moradi, R.; Marto, A.; Rashid, A.S.A.; Moradi, M.M.; Ganiyu, A.A.; Horpibulsuk, S. Bearing Capacity of Soft Soil Model Treated with End-Bearing Bottom Ash Columns. *Environ. Earth Sci.* **2018**, 77, 100. [CrossRef]
- 78. Mah, C.M.; Fujiwara, T.; Ho, C.S. Concrete Waste Management Decision Analysis Based on Life Cycle Assessment. *Chem. Eng. Trans.* **2017**, *56*, 25–30.
- 79. Manning, D.A.C.; Renforth, P.; Lopez-Capel, E.; Robertson, S.; Ghazireh, N. Carbonate Precipitation in Artificial Soils Produced from Basaltic Quarry Fines and Composts: An opportunity for Passive Carbon Sequestration. *Int. J. Greenh. Gas Control* **2013**, *17*, 309–317. [CrossRef]

80. Iszaynuddin, A.H.; Yusairi, B.; Joanes, M. *Industrial Mineral Production Statistics and Directory of Producers in Malaysia*; Mineral Economics Section for Mineral and Geoscience Department Malaysia: Kuala Lumpur, Malaysia, 2017; pp. 2–3.

- 81. Kim, H.-J.; Lee, H.-K. Mineral Sequestration of Carbon Dioxide in Circulating Fluidized Bed Combustion Boiler Bottom Ash. *Minerals* **2017**, *7*, 237. [CrossRef]
- 82. Kolosz, B.; Athanasiadis, I.; Cadisch, G.; Dawson, T.; Giupponi, C.; Honzak, M.; Martinez-Lopez, J.; Marvuglia, A.; Mojtahed, V.; Ogutu, K. Conceptual advancement of socio-ecological modelling of ecosystem services for re-evaluating Brownfield land. *Ecosyst. Serv.* 2018, *33*, 29–39. [CrossRef]
- 83. Pasquier, L.-C.; Mercier, G.; Blais, J.-F.; Cecchi, E.; Kentish, S. Reaction mechanism for the aqueous-phase mineral carbonation of heat-activated serpentine at low temperatures and pressures in flue gas conditions. *Environ. Sci. Technol.* **2014**, *48*, 5163–5170. [CrossRef] [PubMed]



© 2018 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).