

# Pure

Scotland's Rural College

## Diazotroph abundance and community structure are reshaped by straw return and mineral fertilizer in rice-rice-green manure rotation

Yang, L; Bai, J; Zeng, N; Zhou, Xing; Liao, Yulin; Lu, Yanhong; Rees, RM; Nie, Jun; Cao, W

*Published in:*  
Applied Soil Ecology

*DOI:*  
[10.1016/j.apsoil.2018.12.015](https://doi.org/10.1016/j.apsoil.2018.12.015)

First published: 26/12/2018

*Document Version*  
Peer reviewed version

[Link to publication](#)

### *Citation for published version (APA):*

Yang, L., Bai, J., Zeng, N., Zhou, X., Liao, Y., Lu, Y., Rees, RM., Nie, J., & Cao, W. (2018). Diazotroph abundance and community structure are reshaped by straw return and mineral fertilizer in rice-rice-green manure rotation. *Applied Soil Ecology*, 136, 11-20. <https://doi.org/10.1016/j.apsoil.2018.12.015>

### **General rights**

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal ?

### **Take down policy**

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

1 **Diazotroph abundance and community structure are reshaped by rice straw**  
2 **incorporation and inorganic fertilization in rice-rice-green manure rotation**

3

4 Lu Yang <sup>a,b</sup>, Jinshun Bai <sup>b</sup>, Naohua Zeng <sup>b</sup>, Xing Zhou <sup>c,d</sup>, Yulin Liao <sup>d,e</sup>, Yanhong Lu <sup>d,e</sup>,  
5 Robert M. Rees <sup>f</sup>, Jun Nie <sup>d,e,\*</sup>, Weidong Cao <sup>b,\*</sup>

6

7 <sup>a</sup> Graduate School, Chinese Academy of Agricultural Sciences, Beijing 100081, P.R. China

8 <sup>b</sup> Key Laboratory of Plant Nutrition and Fertilizer, Ministry of Agriculture and Rural Affairs /  
9 Institute of Agricultural Resources and Regional Planning, Chinese Academy of Agricultural  
10 Sciences, Beijing 100081, P.R. China

11 <sup>c</sup> College of Resources and Environment, Hunan Agricultural University, Changsha 410128,  
12 P.R. China

13 <sup>d</sup> Soil and Fertilizer Institute of Hunan Province, Hunan Academy of Agricultural Sciences,  
14 Changsha 410125, P.R. China

15 <sup>e</sup> Scientific Observing and Experimental Station of Arable Land Conservation (Hunan),  
16 Ministry of Agriculture and Rural Affairs, Changsha 410125, P.R. China

17 <sup>f</sup> Scotland's Rural College (SRUC), West Mains Road, Edinburgh EH9 3JG, UK

18

19 \* Corresponding author (W.C., [caoweidong@caas.cn](mailto:caoweidong@caas.cn); J.N., [junnie@foxmail.com](mailto:junnie@foxmail.com)).

20

21 **Abstract:**

22 Soil microbial community can be radically changed by the addition of inorganic or organic  
23 nutrients; however, the response of the diazotroph population to nutrient and residue  
24 management in wetland rice rotations is poorly understood. Here, we investigated the  
25 diazotrophic community to inputs of a leguminous green manure Chinese milk vetch (Mv,  
26 *Astragalus sinicus*) using quantitative PCR and Illumina Miseq sequencing of the *nifH* gene.  
27 Five treatments were compared in a Milk vetch–early rice–late rice rotation: 1) Control, no  
28 fertilization and straw return (CK); 2) rice straw incorporation alone (Rs); 3) inorganic  
29 fertilizer application alone (F); 4) F plus Rs (FRs); 5) similar with FRs but high-stubble (~35-  
30 40 cm) of late rice straw was remained (FRh). The results showed that cultivation practice  
31 affected the magnitude of soil available nutrient pools (N, P and K), but not the soil organic  
32 matter and total N pools. The rice straw significantly repressed the *nifH* gene abundance  
33 compared to the control, and also increased the number of diazotrophic bacteria species,  
34 Chao 1 values, Shannon and Simpson indexes more than other fertilization treatments.  
35 Multivariate regression tree analysis revealed that the community diversity and structure of  
36 diazotrophs were primarily shaped by soil nitrate and available P status, as well as C/N ratio.  
37 The most abundant genus *Bradyrhizobium* (21%-32%) tended to decrease in rice straw soil in  
38 comparison with the control but was significantly enhanced in the FRs treatment at seedling  
39 stage and in FRh treatment at flowering stage. Spearman correlation analysis showed that the  
40 dominant diazotrophic genera were positively related with soil available phosphorus, but  
41 responded differentially to soil total N, nitrate, and pH between the seedling and flowering  
42 stages. Overall, the planting and incorporation of vetch under different practices of residue  
43 and fertilizer management reshaped the diazotrophic community during the green manure

44 season, highlighting the crucial roles of soil C, N, and P status or their ratios in shaping the  
45 population and diversity of diazotrophs.

46 **Key words:** Chinese milk vetch; diazotrophic community; inorganic fertilizer; *nifH*, rice;  
47 straw return.

48

## 49 **1. Introduction**

50 The excessive application of inorganic fertilizer nitrogen (N) in agriculture has caused  
51 severe environmental issues (Le et al., 2010; Liu et al., 2013; Gao et al., 2018). A major  
52 challenge for the sustainable production of rice is to maintain grain yield but with lower  
53 environmental impacts (Chen et al., 2014). Substitution of the inorganic fertilizer by organic  
54 substrates, such as straw and green manure is considered a promising way of reducing the  
55 chemical fertilizer input (YadvinderSingh et al., 2004; Xie et al., 2016). Legumes are favored  
56 crops cultivated in the rice-green manure rotation systems because of their ability to fix N by  
57 biological fixation, and enhance N use efficiency and rice productivity (Gao et al., 2011; Xie  
58 et al., 2016). It is estimated that the annual global amount of biological N fixation is 40-70 Tg  
59 in agricultural systems (Galloway et al., 2008; Herridge et al., 2008). Soil microbes are the  
60 key players involved in these processes, and have been well investigated during rice growth  
61 (Wang et al., 2017; Zhang et al., 2017), but little is known about their response to residue  
62 additions in green manure season.

63 The diazotrophs are the major microbial group involved in biological N fixation in rice  
64 paddies, and are highly sensitive to environmental conditions (Jangid et al., 2008), for  
65 instance, the changes of soil physicochemical properties and bio-environment induced by  
66 fertilization and straw incorporation (Bandyopadhyay et al., 2010; Wang et al., 2017).

67 However, there is considerable debate as to how the diazotrophs respond to varied  
68 fertilization practices. Some studies found that the long-term inorganic fertilization affected  
69 both diazotrophic abundance and community composition (Wang et al., 2017), while others  
70 showed marginal effects of fertilization on the abundance of *nifH*-containing microbes  
71 (Mårtensson et al., 2009). This suggests that the abundance and community structure of  
72 diazotrophs could be strongly but differentially affected by inorganic fertilization. The  
73 addition of organic substrates into soil generally has positive influences on diazotrophs. Since  
74 many of the microorganisms participating in N<sub>2</sub> fixation are heterotrophic or mixotrophic, the  
75 addition of external organic matter provides a source of energy and nutrients to support  
76 growth (Rahav et al., 2016). For example, the abundance of the *nifH* gene is enhanced by  
77 stubble retention or straw incorporation from rice plants (Wakelin et al., 2009; Tang et al.,  
78 2017). Moreover, different substrate quality can change soil physicochemical properties, such  
79 as pH, nutrient availability, and the quantity and quality of carbon, which would impose  
80 positive or negative influences on soil microbial populations (Geisseler and Scow, 2014;  
81 Levy-Booth et al., 2014). Besides the exogenous application of inorganic and organic  
82 substrates, plants could be another factor affecting the functions and activities of the soil  
83 microbiome, especially in the rhizosphere (Berendsen et al., 2012). The *nifH* gene abundance  
84 differed between soybean season and wheat season (Sun et al., 2015), implying that the  
85 microbial community may vary with the cultivated crops (Zhang et al., 2017). Thus, it is  
86 important to know how the diazotrophic community changes during the green manure season  
87 within rice-green manure rotations.

88 Chinese milk vetch (Mv) (*Astragalus sinicus* L.) is commonly grown within double-rice  
89 rotations in southern China, where the Mv is planted as winter green manure instead of a bare

90 fallow, and is then incorporated into soil together with rice straw or alone. The returned  
91 leguminous Mv residue is a rich source of N and could stimulate the utilization of rice straw  
92 by the microbial community. Previous studies have shown that the co-incorporation of Mv  
93 and rice straw could change microbial community composition and structure in a paddy soil  
94 (Xie et al., 2017; Gao et al., 2018) or benefit carbon sequestration and improvements in soil  
95 fertility (Poeplau, 2015). It is likely that microbial utilization of C and N within residues is  
96 influenced by the straw return strategy, which includes conventional return (straw crushed  
97 and directly incorporated into soil), straw mulching (straw cover on the field), or high stubble  
98 retention. In practice, straw management is currently changing from conventional return to  
99 high stubble retention because of the popularity of mechanized agricultural operations, which  
100 favor reduced tillage, and labor-saving approaches. To date, many studies have focused on  
101 the influences of green manure incorporation on rice productivity, and microbial or  
102 diazotrophic community changes during the rice growing seasons (Wang et al., 2017; Xie et  
103 al., 2017; Zhang et al., 2017), but little is known about changes in the green manure season,  
104 how the different straw return strategies and their combination with inorganic fertilizer N  
105 application affect the diazotrophic community composition and structure.

106 The study was carried out in a rice-rice-green manure rotation, in which the objectives  
107 were to investigate the influence of legumes, rice straw, and inorganic fertilization on  
108 microbial community composition and structure. Specifically, the two aims were, (i) to  
109 investigate whether different combinations of rice straw incorporation strategies and  
110 inorganic fertilizer application could shift the diazotrophic community composition and  
111 structure in the green manure season; (ii) to identify the key soil physicochemical factors  
112 shaping diazotroph diversity and community composition as affected by the diverse patterns

113 of straw incorporation and inorganic fertilization.

114

## 115 **2. Materials and methods**

### 116 *2.1 Site description and experimental design*

117 The field experiment was carried out from 2011 to 2017 at the Nan County, Hunan Province,  
118 China (29°13'N, 112°28'E) and was conducted on an Entisols, Fluvents, soil derived from  
119 lake sediment with a silt loam texture (USDA soil taxonomy). The basic physicochemical  
120 properties of 0-20 cm soil layer in 2011 were as follows: pH (H<sub>2</sub>O) 7.70, soil organic matter  
121 47.5 g kg<sup>-1</sup>, total N 3.28 g kg<sup>-1</sup>, total P 1.28 g kg<sup>-1</sup>, total K 22.2 g kg<sup>-1</sup>, alkali-hydrolyzale N  
122 261.0 mg kg<sup>-1</sup>, available phosphorus 15.6 mg kg<sup>-1</sup> and available potassium 98.0 mg kg<sup>-1</sup>.

123 The cropping system was early rice, followed by late rice, followed by a winter green  
124 manure within an annual rotation. The early rice (cv. Yuanzao 1) and late rice (cv.  
125 Huanghuazhan) were transplanted in mid-to-late April and mid-to-late July each year,  
126 respectively. The winter green manure Mv was planted at a seed rate of 30 kg ha<sup>-1</sup> in early  
127 October each year and grew through the interval between the late and early rice. Five  
128 treatments were compared in the present study, and Mv crops were incorporated. Beyond  
129 that, each treatment included either an inorganic fertilizer application or rice straw  
130 incorporation, and could be summarized as follows: (i) CK, no fertilizer application and straw  
131 incorporation during the rice-rice-Mv rotation; (ii) Rs, rice straw incorporation alone (both  
132 the early and late rice straw were conventionally returned to the soil after harvest); (iii) F,  
133 inorganic fertilizer application alone (the rates of N, P, and K were 120, 26 and 60 kg ha<sup>-1</sup> for  
134 early rice, and 144, 16 and 80 kg ha<sup>-1</sup> for late rice, respectively); (iv) FRs, F plus Rs; (v) FRh,  
135 the early rice straw was conventionally returned to soil, while high stubble (about 35 cm)

136 from late rice straw was retained after harvest, and returned together with Mv incorporation  
137 before next early rice transplanting. Treatments were arranged in a randomized complete  
138 block design with three replicates. The plot size was 20 m<sup>2</sup> (4 m width by 5 m length).

139 In the treatments without rice straw incorporation, the above-ground parts of rice straw  
140 were removed from the plots. For the inorganic fertilization, half amount of the N (urea) and  
141 all P and K (superphosphate and potassium chloride, respectively) were applied as a base  
142 fertilizer to 5 cm soil depth, and the remaining N was top-dressed at the tillering stage of each  
143 rice season. The nutrients added by inorganic fertilizer or biological N fixation of Mv plants  
144 in each year are listed in Table A1. No fertilizer was applied during Mv's growth.

145

## 146 *2.2 Soil sampling and analysis*

147 Soil samples in the 0-20 cm layer were collected at the Mv seedling and flowering stages  
148 during the green manure season in 2017. Five randomized auger points in each plot were  
149 pooled and mixed thoroughly to provide one sample which was then divided into two sub-  
150 samples and transported to the laboratory on ice. One group was stored at -80°C for DNA  
151 extraction, and the other one was used for soil chemical analysis. The soil ammonium (NH<sub>4</sub><sup>+</sup>)  
152 and nitrate (NO<sub>3</sub><sup>-</sup>) were extracted from fresh soils in 100 ml 0.01 M CaCl<sub>2</sub>, and determined by  
153 continuous flow analysis (Seal AA3, Norderstedt, Germany). The air-dried soil was used for  
154 analysis of other properties as follows. Soil pH was determined with a glass electrode pH  
155 meter (soil:water=1:1, w:v); total N was analyzed by Kjeldahl digestion (Bremner, 1960);  
156 soil organic matter (SOM) was measured using a titration method after oxidation with  
157 K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> (Yeomans and Bremner, 1988); soil available phosphorus (AP) was extracted with  
158 0.5 mol L<sup>-1</sup> NaHCO<sub>3</sub> and analyzed colorimetrically (Murphy and Riley, 1962); soil available



159 potassium (AK) was analyzed by flame photometry following extraction with 1 M  
160 ammonium acetate (Walker and Barber, 1962).

161

### 162 *2.3 DNA extraction, PCR-amplification, and Illumina Miseq sequencing*

163 The genomic DNA was extracted using a PowerSoil® DNA Isolation kit (MO BIO  
164 Laboratories, Inc., CA, USA) following the manufacturer's instructions. The DNA quantity  
165 and quality were assessed using agarose gel electrophoretic analysis and a Nanodrop ND-  
166 1000 spectrophotometer (Nano-Drop Technologies Inc., Wilmington, DE).

167 The primers used for PCR amplification were *nifH* Pol F (5'-TGC GAI CCS AAI GCI  
168 GAC TC-3') and *nifH* AQER (5'-GAC GAT GTA GAT YTC CTG-3') (Poly et al., 2001;  
169 Wartainen et al., 2008). A unique barcode was added at the 5'-end of the reverse primer for  
170 each sample. The PCR amplification program included initial denaturation at 95°C for 3 min,  
171 followed by 30 cycles of 94 °C for 60 s, 56 °C for 60 s, and 72 °C for 60 s, and a final  
172 extension at 72 °C for 10 min. Each sample had two PCR reactions and they were combined  
173 together after amplification. The correct band was excised and purified using a SanPrep DNA  
174 Gel Extraction Kit (Cat# SK8132, Sangon Biotech, Shanghai, China), and was quantified  
175 with a Nanodrop ND-1000 Spectrophotometer. All samples were mixed together in equal  
176 molar amounts from each sample for library construction using a TruSeq DNA kit according  
177 to manufacturer's instructions, and then sequenced by an Illumina Miseq system.

178

### 179 *2.4 Bioinformatic analysis*

180 A total of 28141 sequences were obtained from the 30 samples. The raw sequences were  
181 processed with QIIME pipeline (Caporaso et al., 2010). Low quality sequences were

182 removed, and the barcode was examined to assign sequences to each sample. The chimera  
183 sequences were removed by USEARCH v9.2, and frame shifts were corrected with FrameBot  
184 at default settings. The remaining sequences after quality control were translated into amino  
185 acid sequences for further analysis.

186 The sequences were clustered into operational taxonomic units (OTUs) at the 94%  
187 identity level (Tu et al., 2016). The representative sequences of OTUs were obtained with the  
188 most frequently ones and excluded singletons. Annotation for *nifH* OTUs at an 80% cutoff  
189 value was achieved with reference to the ARB database  
190 ([http://www.zehr.pmc.ucsc.edu/nifH\\_Database\\_Public/](http://www.zehr.pmc.ucsc.edu/nifH_Database_Public/)) (Zehr et al., 2003).

191

## 192 *2.5 Statistical analysis*

193 Data were subjected to analysis of variance using Proc ANOVA with SAS package 9.1 (SAS  
194 Institute, Cary, NC, USA). The cultivation practice (referring to the straw return strategy and  
195 inorganic fertilization) and sampling stage were treated as fixed effects and replication as a  
196 random effect for the data of soil physicochemical properties, gene copy numbers, alpha-  
197 diversity, and relative abundance of different genera. The least significant difference was  
198 used to determine treatment differences at a  $P < 0.05$  level of probability. Spearman  
199 correlation coefficients were employed to test the relationships between soil variables and  
200 relative abundance of genera in R. The principal coordinate analysis (PCoA) was performed  
201 based on the Bray-Curtis distance matrices, and the hierarchy clustering analysis was based  
202 on the weighted unifracs distance. A redundancy analysis (RDA) was conducted to investigate  
203 the correlation between the diazotrophic community composition and soil variables. PCoA,  
204 RDA, and clustering analysis were performed using the ‘vegan’ and ‘GUniFrac’ packages in

205 R. A multivariate regression tree analysis was performed to identify the most important soil  
206 factors for diazotroph diversity and community composition using ‘mvpart’ package (De'Ath,  
207 2002).

208

### 209 **3. Results**

#### 210 *3.1 Effects of straw return and inorganic fertilization on soil properties*

211 The sampling stage and/or cultivation practice had significant influences on the contents of  
212 soil available nutrients ( $\text{NH}_4^+\text{-N}$ ,  $\text{NO}_3^-\text{-N}$ , available P, and available K), soil pH, and C/N  
213 ratio, but not on the SOM and total N contents. No obvious interactions were found between  
214 sampling stage and cultivation practice (Table 1 and Table A2).

215 At the time of Mv planting and incorporation, the Rs treatment led to a marked increase  
216 of soil available K at both seedling and flowering stages of Mv growth, together with an  
217 obvious decrease of  $\text{NO}_3^-\text{-N}$  content at the flowering stage ( $P<0.05$ ; Table 1). Compared to the  
218 Rs treatment, the inorganic fertilizer treatment F resulted in 0.6- to 1.5-fold increase in soil  
219 available P content, but 11-13% less available K, at the two sampling stages. In the  
220 combination treatments (FRs and FRh), there was a further increase of soil available P and K  
221 relative to the Rs and F treatments ( $P<0.05$ ). The soil C/N ratio in these combined treatments  
222 decreased by comparison with the CK at flowering stage, while the pH values increased  
223 ( $P<0.05$ ) (Table 1). There were no significant differences in the soil properties of the FRs and  
224 FRh treatments, except for the lower available K in FRh at the seedling stage (Table 1).

225

#### 226 *3.2 nifH gene copy number*

227 The *nifH* gene copy number ranged from  $4.5 \times 10^6$  to  $25.0 \times 10^6$   $\text{g}^{-1}$  soil across treatments at the

228 Mv seedling stage, which was higher than that at Mv flowering stage ( $11.2 \times 10^6$  to  $15.3 \times 10^6$   
229  $\text{g}^{-1}$  soil) (Fig. 1). At the seedling stage, the gene copy number in the Rs treatment decreased  
230 strongly compared to the CK. The reverse impacts were observed in the inorganic fertilizer  
231 application or by the combinations of FRs and FRh (Fig. 1). At the flowering stage, the  
232 cultivation practice had limited effects on *nifH* gene copy number compared with the CK.  
233 Similar to the pattern at the seedling stage, the gene copy number tended to decrease in the  
234 straw return treatment (Rs), but increased in the combination treatment (FRs), and significant  
235 differences were detected between the Rs and FRs (Fig. 1). The relative influence of soil  
236 variables on *nifH* gene copy number was evaluated by relative weight analysis (Fig. 2).  
237 Nitrate and available P explained most of the variation of *nifH* gene abundance at the  
238 seedling stage (37% and 24%, respectively), while available K and available P were the most  
239 two variables at the flowering stage (27% and 25%, respectively) (Fig. 2)

240

### 241 *3.3 The richness and diversity of diazotrophic community*

242 On the basis of OTUs at 94% similarity, the alpha-diversity of the diazotrophic community  
243 was mainly affected by the cultivation practice, rather than sampling stage, but interactions  
244 were pronounced (Table 2). Greater OTU richness (Chao1 and Observed species) was  
245 observed in the CK and Rs treatments than that in other F-containing treatments where the  
246 inorganic fertilizer tended to repress it. The trends of diazotroph diversity (Shannon diversity  
247 and Simpson index) were consistent with the OTU richness, which showed that the  
248 diazotrophic bacteria in the Rs were more diverse than that in the treatments applied with  
249 inorganic fertilizer (Table 2). Multivariate regression tree analysis indicated that diazotroph  
250 diversity and richness were mainly shaped by soil available P at the seedling stage and by

251  $\text{NO}_3^-$  at the flowering stage (Fig. 3).

252

### 253 *3.4 Diazotrophic community structure*

254 For community composition discrimination, a PCoA was performed based on the Bray-  
255 Curtis distance of OTU compositions (Fig. 4a and b). At the seedling stage, PCoA analysis  
256 showed that Rs was separated from other groups at both seedling and flowering stages of Mv  
257 growth. The hierarchy clustering analysis further confirmed these results, and the biggest  
258 differences in diazotroph structure occurred between Rs and FRs at seedling and between Rs  
259 and FRh at flowering stage, respectively (Fig. 4 c and d). In addition, the diazotrophic  
260 communities formed four clusters at the Mv flowering: CK, Rs, FRh, and others, showing  
261 that diazotroph communities differed between treatments (Fig. 4 b and d).

262 The redundancy analysis showed that the examined soil variables explained 50.9% of the  
263 variation of diazotroph community composition at the Mv seedling stage, and the first two  
264 axes attributed to 19.6% and 9.3%, respectively (Fig. 5a). According to the vectors, the  
265 diazotrophic communities of the Rs treatment were positively correlated with soil C/N ratio  
266 and pH, whereas in the FRs and FRh treatments, the communities were associated with soil  
267  $\text{NO}_3^-$ , available P, total N, and SOM contents (Fig. 5a). At the Mv flowering stage, soil  
268 variables explained 74.7% of the variation, which was attributed to soil TN ( $F=2.71$ ,  
269  $P=0.049$ ), SOM ( $F=2.93$ ,  $P=0.029$ ), available P ( $F=3.34$ ,  $P=0.007$ ), and available K ( $F=2.36$ ,  
270  $P=0.023$ ) (Fig. 5b). The diazotrophic community of the Rs treatment was associated with  
271 SOM, soil total N, and the C:N ratio, while it was affected more by soil available K in FRs  
272 treatment, and by soil pH, available P and  $\text{NH}_4^+$  in the FRh treatment (Fig. 5b).

273

### 274 3.5 Relative abundance of the diazotrophic taxa

275 OTUS of all treatment samples were taxonomically classified into 14 different genera which  
276 accounted for about 70% of the total sequences, with the other unclassified genera attributed  
277 to the remaining 30%. The seven most abundant genera with relative abundance more than  
278 1% are presented in Fig. 6. The genus *Bradyrhizobium* was by far the most dominant group in  
279 all treatments, containing 21.8%-29.1% and 21.4%-32.2% of the total *nifH* gene sequences in  
280 soils at the Mv seedling and flowering stages, respectively (Fig. 6). The relative abundance of  
281 *Bradyrhizobium* was slightly reduced in the Rs soil by comparison with the CK, but was  
282 significantly increased ( $P<0.05$ ) by FRs at seedling stage and by FRh at flowering stage. The  
283 *Anaeromyxobacter* and *Burkholderia* abundances were significantly increased in the Rs  
284 treatment by comparison with the CK at the flowering stage. There were no differences in the  
285 genus abundance between the FRs and FRh treatments, except for greater abundance of  
286 *Anaeromyxobacter* in FRh at seedling stage (Fig. 6). Multivariate regression tree analysis  
287 showed that the diazotroph community composition was mainly shaped by available P and  
288 available K at the Mv seedling stage and by C/N ratio at the flowering stage, respectively  
289 (Fig. 7).

290 A Spearman analysis indicated that soil variables were closely correlated, but  
291 differentially, with most of the abundant genera (Fig. 8). As to the most abundant two genera,  
292 the relative abundance of *Bradyrhizobium* at the Mv flowering stage was closely related to  
293 soil available P and negatively to the C/N ratio. The *Azospirillum* abundance was positively  
294 related to soil  $\text{NO}_3^-$  and available P contents (Fig. 8). By contrast with *Bradyrhizobium* and  
295 *Azospirillum*, the relative abundance of *Rhodobacter* was closely associated with SOM, TN,  
296 available P and negatively correlated with pH at the seedling stage. In addition, the relative

297 abundance of *Anaeromyxobacter* and *Burkholderia* genera were negatively correlated with  
298 soil NO<sub>3</sub><sup>-</sup> content at seedling stage and total N at flowering stage, respectively.

299

## 300 **4. Discussion**

301

302 *4.1 Successive straw return and inorganic fertilizer application changed the diazotrophic*  
303 *abundance and richness in green manure season*

304 Soil microbes are highly sensitive to the environmental changes, such as soil C, N, and P  
305 availability or their ratios. It has been reported that long-term inorganic fertilizer application  
306 generally decreases soil microbial abundance (Wang et al., 2017), while the frequent return of  
307 rice straw could increase the diazotrophic diversity and richness, but reduce the *nifH* gene  
308 expression (Tang et al., 2017). Similar results were found in the present study where the *nifH*  
309 gene copy number was increased (relative to the control) in the treatments with inorganic  
310 fertilizer application but decreased in the treatment of straw incorporation alone (Fig. 1). The  
311 diazotrophic community was more diverse in the straw treatment, and less so in the  
312 treatments receiving inorganic fertilizer (Table 2). The results indicated that the diazotroph  
313 communities were altered by the straw and fertilization practices.

314 Nitrogen and P are essential for the growth and metabolism of organisms and play a  
315 pivotal role in biological N fixation (Wurzburger et al., 2012). Depending on the various  
316 environmental factors in different ecosystems, N availability may have contrasting effects on  
317 the abundance of *nifH* gene, where enhanced and suppressed impacts have been both  
318 documented (Gonzalez Perez et al., 2014; Reardon et al., 2014; Zhalnina et al., 2015; Wang  
319 et al., 2017). Nonetheless, high soil NO<sub>3</sub><sup>-</sup>-N and NH<sub>4</sub><sup>+</sup>-N content generally inhibit the

320 members of the diazotrophic community, while lower N status could stimulate the biological  
321 N fixation (Coelho et al., 2008; Wang et al., 2017). Rice straw has a relatively low N  
322 concentration, but is rich in C. In the Rs treatment, rice straw incorporation might be  
323 expected to stimulate microbial growth and enhance the immobilization of soil available N  
324 (Rao and Mikkelsen, 1976). The lower soil NO<sub>3</sub><sup>-</sup>-N content was observed in Rs treatment is  
325 consistent with this expectation (Table 1). Furthermore, the significantly lower soil available  
326 P content in the Rs treatment compared with that in the F or combination treatments might  
327 further exacerbate such impacts. In agreement with that, soil nitrate and available P were  
328 identified as the primary predictor of *nifH* gene abundance (Fig. 2). In addition, the increased  
329 *nifH* gene abundance in the Rs treatment might be partly attributed to the increased relative  
330 abundance of dominant diazotroph genera *Anaeromyxobacter* and *Burkholderia* (Fig. 6).

331 The results also showed that diversity was closely related to soil C/N ratios. Higher C/N  
332 values could result in a competitive advantage for free living diazotrophs (Mirza et al., 2014).  
333 Due to the relatively less N input, the CK and Rs treatments had greater C/N ratios by  
334 comparison with the F, FRs, and FRh treatments at the Mv flowering stage, which might also  
335 explain the higher diversity in the Rs treatment than others. Consistently, the diazotroph  
336 diversity at this stage was mainly shaped by soil NO<sub>3</sub><sup>-</sup> content and C/N ratios (Fig. 3b). Taken  
337 together, the gene copy numbers and diazotrophic diversity had relatively moderate values in  
338 the combinations of fertilization and rice straw incorporation, suggesting that both the N and  
339 P nutrients and C supply are important factors in regulating the shifts of diazotrophic  
340 diversity and richness.

341 Soil pH has been considered as a key factor affecting soil microbial richness and  
342 diversity (Levy-Booth et al., 2014; Wang et al., 2017). It is often due to the negative feedback



343 from intensified soil acidification caused by long-term inorganic fertilization (Guo et al.,  
344 2010). This could also partly explain the higher diversity in Rs treatment relative to those  
345 receiving fertilizer additions since higher pH values were observed in the straw treatment  
346 than the control. However, the reduction in the diversity index was not been fully offset by  
347 the combinations of fertilization and straw incorporation, which might imply the strong  
348 impacts of long-term inorganic fertilizer application on the diversity of the diazotrophic  
349 community.

350

#### 351 *4.2 Straw return and inorganic fertilizer application reshaped the diazotrophic community* 352 *structure during the growth of Chinese milk vetch*

353 Straw mulching and return, especially when combined with inorganic fertilization, could  
354 significantly reshape the soil diazotrophic community structure (Tang et al., 2017). Many  
355 studies have found that the community structure of soil ammonia-oxidizing bacteria,  
356 ammonia-oxidizing archaea, fungi, and diazotrophic bacteria could be easily affected by the  
357 long-term fertilization (Geisseler and Scow, 2014). In this study, the shifts of diazotrophic  
358 community structure were consistently observed in response to straw mulching and fertilizer  
359 application (Fig. 4). Compared to the control, the diazotrophic community structure under the  
360 Rs treatment was separated from others during the Mv growth, and the biggest differences in  
361 diazotroph structure occurred between Rs and FRs at seedling and between Rs and FRh at  
362 flowering stage, respectively (Fig. 4), which showed that cultivation practices indeed affected  
363 the soil diazotrophic community structure.

364 Soil is a very complex environment and the formation of microbial community's  
365 structure is often regulated by various interacting soil variables (Young, 1998; Geisseler and

366 Scow, 2014). The availability of soil N, P, and K are the important factors shaping  
367 diazotrophic community composition (Tang et al., 2017; Wang et al., 2017). Varied  
368 responses of the diazotrophic community structure and populations to soil nutrient status  
369 were observed, which might have resulted from the greater availability of soil  $\text{NO}_3^-$  and  
370 available P in the treatments with inorganic fertilizer application relative to the Rs treatment.  
371 In accordance with that, the diazotrophic communities in the FRs and FRh treatments were  
372 associated with soil  $\text{NO}_3^-$ , available P, total N, and SOM contents at the Mv seedling stage,  
373 and affected more by soil available K in FRs treatment, and by soil pH, available P and  $\text{NH}_4^+$   
374 in the FRh treatment (Fig. 5). The results indicated that soil nutrient availability, which is  
375 highly responsive to fertilizer input, is crucial for the establishment of soil diazotrophic  
376 community structure.

377 Likewise, our results also showed that the soil C/N ratio and soil organic matter  
378 concentration played important roles in regulating soil diazotrophic community composition.  
379 The quality and quantity of added straw could influence the response to different fertilization  
380 practices by altering microbial mineralization and immobilization turnover (Curtin and Fraser,  
381 2003; Cusack et al., 2011). However, the practice of straw incorporation, i.e. conventionally  
382 returned to soil and high stubble retention, under the present conditions, had marginal  
383 differences in shaping the diazotrophic community and structure.

384

#### 385 *4.3 Straw return and inorganic fertilizer application influenced the dominant diazotrophic* 386 *genera during the growth of Chinese milk vetch*

387 The diazotrophic genera may respond differently to the various soil physicochemical  
388 properties, due to their individual sensitivities to changes of the soil environment (Wakelin et

389 al., 2009; Wang et al., 2017). There were clearly various responses or sensitivities of the  
390 different dominant diazotrophic genera to soil conditions in this study (Figs. 6, 7 and 8).  
391 Cultivation practices such as fertilization and straw incorporation are common ways of  
392 altering environmental conditions of soil microbes, and thus could change the composition of  
393 diazotrophic genera depending on their competitiveness and external environment. The  
394 *Bradyrhizobium* genus is ubiquitous in soil and is a commonly known N<sub>2</sub>-fixing bacterium,  
395 which includes symbiotic N-fixing bacteria and free-living soil diazotrophs (Kahindi et al.,  
396 1997). As expected, the genus *Bradyrhizobium* was the most dominant group in all  
397 treatments, containing 21%-32% of the total *nifH* gene sequences in soils at the seedling and  
398 flowering stages of Chinese milk vetch (Fig. 6). The relative abundance of *Bradyrhizobium*  
399 was slightly reduced in Rs soil in comparison with CK, but significantly increased in FRs at  
400 seedling stage and in FRh at flowering stage. This was likely to have caused by the lower  
401 level of soil available P and higher C/N ratio in Rs treatment, since the relative abundance of  
402 *Badyrhizobium* was positively related to the soil's available P content, but negatively  
403 correlated with soil C/N ratios (Fig. 8). These results were in agreement with the concept that  
404 the dominant genera abundance was mainly shaped by the soil available P and C/N ratio (Fig.  
405 7). The significant correlations between the relative abundance of *Azospirillum* and  
406 *Rhodobacter* and the content of soil available N and P implied that these two diazotrophic  
407 genera were very sensitive to soil N and P nutrients status. Interestingly, the Rs treatment  
408 selectively increased the relative abundance of *Anaeromyxobacter* and *Burkholderia*,  
409 especially at the flowering stage. They were negatively associated with either soil nitrate or  
410 total N content, but positively related to soil C/N and pH. The lower soil nitrate content in Rs  
411 treatment further indicated that these two dominant genera were more sensitive to soil C and

412 N status than others.

413 Changes in soil quality occur slowly and need time for the soil microbial community to  
414 adapt to the altered environments and then stabilize (Geisseler and Scow, 2014; Wang et al.,  
415 2017). In the present study, we found that the relative abundance of the dominant genera  
416 *Bradyrhizobium* was generally lower than that reported by other studies during the rice  
417 growing seasons (Wang et al., 2017). This might be due to the different soil texture and  
418 different sampling seasons. The soil samples were collected during the winter season of the  
419 Chinese milk vetch, which had relatively lower temperatures than the rice seasons, thus  
420 leading to lower overall microbial activity. In addition, the differences might also be  
421 associated with the cultivated plants, since different *nifH* gene abundances between soybean  
422 and wheat seasons have been reported (Sun et al., 2015). Nitrogen fixing bacteria strongly  
423 interact with cultivated plants (Leguminosae vs. Poaceae), which could change the sensitivity  
424 and responses of diazotrophic genera to various soil conditions (Tan et al., 2003).

425 Taken together, at the time of Chinese milk vetch planting and incorporation, straw  
426 incorporation alone had relatively sufficient C inputs, but was short of N and P, which  
427 repressed the *nifH* gene abundance and the relative abundance of dominant diazotroph  
428 genera. The results highlighted the importance of straw incorporation coupled with N and P  
429 inputs to maintain microbial activity and diversity in the rice-rice-green manure rotations in  
430 subtropical regions.

431

## 432 **5. Conclusion**

433 In conclusion, this study has demonstrated that straw incorporation and inorganic fertilizer  
434 application lead to significant changes in soil diazotrophic community structure and

435 populations during the growth of Chinese milk vetch in a paddy soil. The straw alone could  
436 supply enough C to support microbial activity, but it contained relatively little N and P, thus  
437 reducing the *nifH* gene abundance and tending to decrease the relative abundance of the  
438 diazotrophic genera. The high C inputs with lower nutrient availability in straw enhanced  
439 diazotrophic diversity, while successive inorganic fertilizer applications decreased it.  
440 Although the combination of Mv, straw, and fertilization did not fully reverse the decline, the  
441 co-utilization practice helped stimulate the dominant diazotroph abundance by comparison  
442 with straw alone. Overall, the results suggest that both the C sources introduced by the rice  
443 straw and the N and P supplied by fertilizer application were crucial for improving soil  
444 quality and the population and diversity of soil diazotrophic community.

445

#### 446 **Acknowledgements**

447 This work was supported by the China Agriculture Research System - Green Manure; the  
448 Science and Technology Innovation Project of Chinese Academy of Agricultural Sciences;  
449 the Chinese Outstanding Talents Program in Agricultural Science; and the Newton Fund  
450 [Grant Ref: BB/N013484/1].

451 **Declarations of interest:** none.

452 **Appendix A.** Supplementary data

#### 453 **References**

454 Bandyopadhyay, K.K., Misra, A.K., Ghosh, P.K., Hati, K.M., 2010. Effect of integrated use  
455 of farmyard manure and chemical fertilizers on soil physical properties and productivity  
456 of soybean. *Soil Till. Res.* 110, 115-125.  
457 Berendsen, R.L., Pieterse, C.M.J., Bakker, P.A.H.M., 2012. The rhizosphere microbiome and

458 plant health. Trends Plant Sci. 17, 478-486.

459 Bremner, J.M., 1960. Determination of nitrogen in soil by the Kjeldahl method. J. Agric. Sci.  
460 55, 11-33.

461 Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K.,  
462 Fierer, N., Peña, A.G., Goodrich, J.K., Gordon, J.I., Huttley, G.A., Kelley, S.T., Knights,  
463 D., Koenig, J.E., Ley, R.E., Lozupone, C.A., McDonald, D., Muegge, B.D., Pirrung, M.,  
464 Reeder, J., Sevinsky, J.R., Turnbaugh, P.J., Walters, W.A., Widmann, J., Yatsunenko, T.,  
465 Zaneveld, J., Knight, R., 2010. QIIME allows analysis of high-throughput community  
466 sequencing data. Nat. Methods 7, 335.

467 Chen, X., Cui, Z., Fan, M., Vitousek, P., Zhao, M., Ma, W., Wang, Z., Zhang, W., Yan, X.,  
468 Yang, J., Deng, X., Gao, Q., Zhang, Q., Guo, S., Ren, J., Li, S., Ye, Y., Wang, Z., Huang,  
469 J., Tang, Q., Sun, Y., Peng, X., Zhang, J., He, M., Zhu, Y., Xue, J., Wang, G., Wu, L., An,  
470 N., Wu, L., Ma, L., Zhang, W., Zhang, F., 2014. Producing more grain with lower  
471 environmental costs. Nature 514, 486-489.

472 Coelho, M.R.R., Vos, D.M., Carneiro, N.P., Marriel, I.E., Paiva, E., Seldin, L., 2008.  
473 Diversity of nifH gene pools in the rhizosphere of two cultivars of sorghum (*Sorghum*  
474 *bicolor*) treated with contrasting levels of nitrogen fertilizer. FEMS Microbiol. Lett. 279,  
475 15-22.

476 Curtin, D., Fraser, P.M., 2003. Soil organic matter as influenced by straw management  
477 practices and inclusion of grass and clover seed crops in cereal rotations. Soil Res. 41, 95-  
478 106.

479 Cusack, D.F., Silver, W.L., Torn, M.S., Burton, S.D., Firestone, M.K., 2011. Changes in  
480 microbial community characteristics and soil organic matter with nitrogen additions in

481 two tropical forests. *Ecology* 92, 621-632.

482 De'Ath, G., 2002. Multivariate regression trees: a new technique for modeling species -  
483 environment relationships. *Ecology* 83, 1105-1117.

484 Galloway, J.N., Townsend, A.R., Erisman, J.W., Bekunda, M., Cai, Z., Freney, J.R.,  
485 Martinelli, L.A., Seitzinger, S.P., Sutton, M.A., 2008. Transformation of the nitrogen  
486 cycle: recent trends, questions, and potential solutions. *Science* 320, 889-892.

487 Gao, B., Huang, T., Ju, X., Gu, B., Huang, W., Xu, L., Rees, R.M., Powlson, D.S., Smith, P.,  
488 Cui, S., 2018. Chinese cropping systems are a net source of greenhouse gases despite soil  
489 carbon sequestration. *Global Change Biol.* In Press. <https://doi.org/10.1111/gcb.14425>

490 Gao, S., Cao, W., Zou, C., Gao, J., Huang, J., Bai, J., Zeng, N., Shimizu, K., Wright, A., Dou,  
491 F., 2018. Ammonia-oxidizing archaea are more sensitive than ammonia-oxidizing  
492 bacteria to long-term application of green manure in red paddy soil. *Appl. Soil Ecol.* 124,  
493 185-193.

494 Geisseler, D., Scow, K.M., 2014. Long-term effects of mineral fertilizers on soil  
495 microorganisms—A review. *Soil Biol. Biochem.* 75, 54-63.

496 Gonzalez Perez, P., Ye, J., Wang, S., Wang, X., Huang, D., 2014. Analysis of the occurrence  
497 and activity of diazotrophic communities in organic and conventional horticultural soils.  
498 *Appl. Soil Ecol.* 79, 37-48.

499 Guo, J.H., Liu, X.J., Zhang, Y., Shen, J.L., Man, W.X., Zhang, W.F., Christie, P., Goulding,  
500 K.W.T., Vitousek, P.M., Zhang, F.S., 2010. Significant acidification in major Chinese  
501 croplands. *Science* 327, 1008-1010.

502 Herridge, D.F., Peoples, M.B., Boddey, R.M., 2008. Global inputs of biological nitrogen  
503 fixation in agricultural systems. *Plant Soil* 311, 1-18.

504 Jangid, K., Williams, M.A., Franzluebbers, A.J., Sanderlin, J.S., Reeves, J.H., Jenkins, M.B.,  
505 Endale, D.M., Coleman, D.C., Whitman, W.B., 2008. Relative impacts of land-use,  
506 management intensity and fertilization upon soil microbial community structure in  
507 agricultural systems. *Soil Biol. Biochem.* 40, 2843-2853.

508 Kahindi, J.H.P., Woomer, P., George, T., de Souza Moreira, F.M., Karanja, N.K., Giller, K.E.,  
509 1997. Agricultural intensification, soil biodiversity and ecosystem function in the tropics:  
510 the role of nitrogen-fixing bacteria. *Appl. Soil Ecol.* 6, 55-76.

511 Le, C., Zha, Y., Li, Y., Sun, D., Lu, H., Yin, B., 2010. Eutrophication of lake waters in China:  
512 cost, causes, and control. *Environ. Manage.* 45, 662.

513 Levy-Booth, D.J., Prescott, C.E., Grayston, S.J., 2014. Microbial functional genes involved in  
514 nitrogen fixation, nitrification and denitrification in forest ecosystems. *Soil Biol. Biochem.*  
515 75, 11-25.

516 Liu, X., Zhang, Y., Han, W., Tang, A., Shen, J., Cui, Z., Vitousek, P., Erisman, J.W.,  
517 Goulding, K., Christie, P., Fangmeier, A., Zhang, F., 2013. Enhanced nitrogen deposition  
518 over China. *Nature* 494, 459-462.

519 Mårtensson, L., Díez, B., Wartiainen, I., Zheng, W., Elshehawy, R., Rasmussen, U., 2009.  
520 Diazotrophic diversity, *nifH* gene expression and nitrogenase activity in a rice paddy field  
521 in Fujian, China. *Plant Soil* 325, 207-218.

522 Mirza, B.S., Potisap, C., Nüsslein, K., Bohannan, B.J.M., Rodrigues, J.L.M., 2014. Response  
523 of free-living nitrogen-fixing microorganisms to land use change in the amazon rainforest.  
524 *Appl. Environ. Microb.* 80, 281-288.

525 Murphy, J., Riley, J.P., 1962. A modified single solution method for the determination of  
526 phosphate in natural waters. *Anal. Chim. Acta.* 27, 31-36.



527 Poelau, C., 2015. Carbon sequestration in agricultural soils via cultivation of cover crops  
528 A meta-analysis. *Agr. Ecosyst. Environ.* 200, 33-41.

529 Poly, F., Monrozier, L.J., Bally, R., 2001. Improvement in the RFLP procedure for studying  
530 the diversity of nifH genes in communities of nitrogen fixers in soil. *Res. Microbiol.* 152,  
531 95-103.

532 Rahav, E., Giannetto, M.J., Bar-Zeev, E., 2016. Contribution of mono and polysaccharides to  
533 heterotrophic N<sub>2</sub>fixation at the eastern Mediterranean coastline. *Sci. Rep-UK* 6, 27858.

534 Rao, D.N., Mikkelsen, D.S., 1976. Effect of Rice Straw Incorporation on Rice Plant Growth  
535 and Nutrition. *Agron. J.* 68, 752-755.

536 Reardon, C.L., Gollany, H.T., Wuest, S.B., 2014. Diazotroph community structure and  
537 abundance in wheat–fallow and wheat–pea crop rotations. *Soil Biol. Biochem.* 69, 406-  
538 412.

539 Sun, R., Guo, X., Wang, D., Chu, H., 2015. Effects of long-term application of chemical and  
540 organic fertilizers on the abundance of microbial communities involved in the nitrogen  
541 cycle. *Appl. Soil Ecol.* 95, 171-178.

542 Tan, Z., Hurek, T., Reinhold-Hurek, B., 2003. Effect of N-fertilization, plant genotype and  
543 environmental conditions on nifH gene pools in roots of rice. *Environ. Microbiol.* 5,  
544 1009-1015.

545 Tang, Y., Zhang, W., Zhang, M., Chen, A., Wei, W., Sheng, R., 2017. Impact of fertilization  
546 regimes on diazotroph community compositions and N<sub>2</sub>-fixation activity in paddy soil.  
547 *Agr. Ecosyst. Environ.* 247, 1-8.

548 Tu, Q., Zhou, X., He, Z., Xue, K., Wu, L., Reich, P., Hobbie, S., Zhou, J., 2016. The  
549 Diversity and Co-occurrence Patterns of N<sub>2</sub>-Fixing Communities in a CO<sub>2</sub>-Enriched

550 Grassland Ecosystem. *Microb. Ecol.* 71, 604-615.

551 Wakelin, S.A., Gregg, A.L., Simpson, R.J., Li, G.D., Riley, I.T., McKay, A.C., 2009. Pasture  
552 management clearly affects soil microbial community structure and N-cycling bacteria.  
553 *Pedobiologia* 52, 237-251.

554 Walker, J.M., Barber, S.A., 1962. Absorption of potassium and rubidium from the soil by  
555 corn roots. *Plant Soil* 17, 243-259.

556 Wang, C., Wang, B., Zheng, M., Song, W., Wen, S., Zhu, C., Shen, R., 2017. Impact of 25  
557 years of inorganic fertilization on diazotrophic abundance and community structure in an  
558 acidic soil in southern China. *Soil Biol. Biochem.* 113, 240-249.

559 Wang, Y., Wang, C., Wang, J., Li, X., Li, H., Li, C., Kou, Y., Tu, B., Yao, M., 2017. Soil pH  
560 is a major driver of soil diazotrophic community assembly in Qinghai-Tibet alpine  
561 meadows. *Soil Biol. Biochem.* 115, 547-555.

562 Warttinen, I., Eriksson, T., Zheng, W., Rasmussen, U., 2008. Variation in the active  
563 diazotrophic community in rice paddy—nifH PCR-DGGE analysis of rhizosphere and bulk  
564 soil. *Appl. Soil Ecol.* 39, 65-75.

565 Wurzbarger, N., Bellenger, J.P., Kraepiel, A.M.L., Hedin, L.O., 2012. Molybdenum and  
566 phosphorus interact to constrain asymbiotic nitrogen fixation in tropical forests. *PloS one*  
567 7, e33710.

568 Xie, Z., He, Y., Tu, S., Xu, C., Liu, G., Wang, H., Cao, W., Liu, H., 2017. Chinese milk vetch  
569 improves plant growth, development and <sup>15</sup>N recovery in the rice-based rotation system  
570 of south China. *Sci. Rep-UK* 7.

571 Xie, Z., Tu, S., Shah, F., Xu, C., Chen, J., Han, D., Liu, G., Li, H., Muhammad, I., Cao, W.,  
572 2016. Substitution of fertilizer-N by green manure improves the sustainability of yield in

573 double-rice cropping system in south China. *Field Crops Res.* 188, 142-149.

574 YadvinderSingh, BijaySingh, Timsina, J., 2004. Crop residue management for nutrient  
575 cycling and improving soil productivity in rice-based cropping systems in the tropics.  
576 *Adv. Agron.* 85, 269-407.

577 Yeomans, J.C., Bremner, J.M., 1988. A rapid and precise method for routine determination of  
578 organic carbon in soil 1. *Commun. Soil Sci. Pl. An.* 19, 1467-1476.

579 Young, I.M., 1998. Biophysical interactions at the root – soil interface: a review. *J. Agric. Sci.*  
580 130, 1-7.

581 Zehr, J.P., Jenkins, B.D., Short, S.M., Steward, G.F., 2003. Nitrogenase gene diversity and  
582 microbial community structure: a cross-system comparison. *Environ. Microbiol.* 5, 539-  
583 554.

584 Zhalnina, K., Dias, R., de Quadros, P.D., Davis-Richardson, A., Camargo, F.A.O., Clark,  
585 I.M., McGrath, S.P., Hirsch, P.R., Triplett, E.W., 2015. Soil pH Determines Microbial  
586 Diversity and Composition in the Park Grass Experiment. *Microb. Ecol.* 69, 395-406.

587 Zhang, X., Zhang, R., Zhang, C., Gao, J., Wang, X., Fan, F., Ma, X., Yin, H., Feng, K., Deng,  
588 Y., 2017. Thirty-one years of rice-rice-green manure rotations shape the rhizosphere  
589 microbial community and enrich beneficial bacteria. *Soil Biol. Biochem.* 104, 208-217.

590

591 **Tables**592 **Table 1.** Soil physicochemical properties at seedling and flowering stages of Chinese milk vetch.

Stage	Treatment	Soil organic matter (g/kg)	Total N (g/kg)	NO <sub>3</sub> <sup>-</sup> (mg/kg)	NH <sub>4</sub> <sup>+</sup> (mg/kg)	Available P (mg/kg)	Available K (mg/kg)	C/N ratio	pH
Seedling	CK	45.9±0.4 a	2.8±0.0 a	25.7±2.3 b	5.5±1.4 a	11.5±0.8 c	88.2±0.8 d	9.5±0.0 a	7.67±0.01 b
	Rs	44.5±0.4 a	2.7±0.1 a	25.8±0.9 b	5.6±1.3 a	10.1±0.3 c	106.0±1.5 c	9.7±0.1 a	7.73±0.01 a
	F	46.2±1.0 a	2.8±0.0 a	30.0±4.9 ab	4.3±1.6 a	16.3±0.6 b	94.4±0.8 cd	9.7±0.1 a	7.69±0.02 ab
	FRs	44.3±1.0 a	2.8±0.1 a	35.9±2.6 a	3.8±1.7 a	19.7±2.2 ab	139.3±4.7 a	9.3±0.2 a	7.68±0.02 b
	FRh	46.0±0.9 a	2.8±0.0 a	34.2±4.0 ab	5.2±1.3 a	21.2±1.2 a	121.2±6.6 b	9.7±0.2 a	7.67±0.01 b
Flowering	CK	45.7±0.2 a	2.7±0.0 a	4.7±0.8 a	1.4±0.8 a	5.7±0.4 bc	78.9±1.6 c	9.9±0.1 a	7.73±0.02 b
	Rs	46.0±3.3 a	2.7±0.2 a	2.6±0.3 b	1.2±0.6 a	2.8±1.1 c	93.1±3.3 b	10.1±0.1 a	7.79±0.05 ab
	F	43.6±0.3 a	2.6±0.0 a	3.4±0.2 ab	2.7±0.0 a	7.1±0.3 b	81.0±1.3 c	9.7±0.1 ab	7.77±0.04 ab
	FRs	46.8±2.1 a	2.9±0.1 a	3.3±0.3 ab	2.0±0.5 a	12.4±1.4 a	109.7±3.4 a	9.5±0.1 b	7.85±0.03 a
	FRh	43.0±1.2 a	2.6±0.0 a	4.3±0.5 ab	2.4±0.1 a	11.8±1.4 a	105.8±2.4 a	9.5±0.1 b	7.87±0.02 a

593 Values indicate mean±SE (*n*=3). Different letters in the columns represent significant differences between treatments within each stage (*P* < 0.05).

594

595 **Table 2.** The diazotrophic alpha-diversity as influenced by varied straw return and  
 596 fertilization practices in a Chinese milk vetch-based system.

Stage	Treatment	Observed species	Chao1 index	Shannon diversity	Simpson index
Seedling	CK	1623±103 a	2746.0±62.4 a	7.5±0.1 a	0.982±0.002 ab
	Rs	1539±31 ab	2564.0±43.1 ab	7.6±0.1 a	0.982±0.003 ab
	F	1332±32 cd	2312.0±101.5 c	7.2±0.0 bc	0.978±0.001 bc
	FRs	1341±37 cd	2205.0±5.5 c	7.1±0.1 bc	0.978±0.001 bc
	FRh	1448±82 bc	2368.0±137.3 bc	7.4±0.1 ab	0.981±0.001 ab
Flowering	CK	1399±51 bc	2356.3±50.4 bc	7.2±0.1 bc	0.979±0.002 bc
	Rs	1550±56 ab	2672.3±99.9 a	7.6±0.2 a	0.984±0.003 a
	F	1384±34 cd	2299.7±93.4 c	7.2±0.0 bc	0.978±0.000 bc
	FRs	1354±20 cd	2232.3±41.9 c	7.3±0.0 ab	0.981±0.001 ab
	FRh	1244±42 d	2160.7±75.1 c	7.0±0.1 c	0.975±0.001 c
Stage (S)		0.0424	0.0870	0.0597	0.3249
Cultivation practice (C)		0.0010	0.0034	<0.0001	0.0271
S × C		0.0317	0.0320	0.0238	0.0607

597 Values indicate mean±SE ( $n=3$ ). Different letters in the columns represent significant  
 598 differences between treatments across stages ( $P < 0.05$ ).

599

600 **Figure legend**

601

602 **Figure 1. *nifH* gene copy number in soils sampled at the seedling and flowering stages of**  
603 **Chinese milk vetch as affected by the continuous straw return or fertilization practices.**

604 The data were subjected to a two-way analysis of variance. Abbreviations: S, stage; C,  
605 cultivation practice; Different capital letters above the gray columns and lowercase letters  
606 above the black columns indicate significant differences among cultivation practices at  
607 seedling and flowering stages by LSD test, respectively ( $P < 0.05$ ). The asterisks within each  
608 cultivation practice indicate significant differences between two stages (\*,  $P < 0.05$ ; \*\*,  $P <$   
609  $0.01$ ). Vertical bars represent the standard error of four replicates.

610

611 **Figure 2. Relative influence of soil physicochemical properties on abundance of *nifH***  
612 **genes at the seedling and flowering stage of Chinese milk vetch evaluated using relative**  
613 **weight analysis.** Abbreviations: AK, available potassium; AP, available phosphorus; C/N,  
614 carbon to nitrogen ratio; SOM, soil organic matter; TN, total nitrogen. Values are means of  
615 three replicates.

616

617 **Figure 3. Multivariate regression tree analysis of alpha diversity (observed species,**  
618 **Chao 1, Shannon and Simpson indexes) of diazotrophs and soil physicochemical**  
619 **variables.** Treatments and the number of samples included in the analysis are shown at the  
620 bottom. Abbreviations: AK, available potassium; AP, available phosphorus; C/N, carbon to  
621 nitrogen ratio; SOM, soil organic matter; TN, total nitrogen.

622

623 **Figure 4. Principal coordinate analysis of diazotrophic community composition in soils**  
624 **sampled at the seedling and flowering stages of Chinese milk vetch as affected by the**  
625 **continuous straw return or fertilization practices.** The samples were analyzed in triplicate  
626 plots.

627

628 **Figure 5. Redundancy analysis of the diazotrophic community structure in soils**  
629 **sampled at the seedling and flowering stages of Chinese milk vetch as affected by the**  
630 **continuous straw return or fertilization practices.** The positions and lengths of the arrows  
631 indicate the directions and strengths, respectively, of the effects of variables on the  
632 diazotrophic communities. The samples were analyzed in triplicate plots. Abbreviations: AK,  
633 available potassium; AP, available phosphorus; C/N, carbon to nitrogen ratio; SOM, soil  
634 organic matter; TN, total nitrogen.

635

636 **Figure 6. Relative abundances (%) of the seven most abundant genera (>1%) in soils**  
637 **sampled at the seedling and flowering stages of Chinese milk vetch as affected by the**  
638 **continuous straw return or fertilization practices.** The samples were analyzed in triplicate  
639 plots. Different letters above columns in each genus indicate significant differences among  
640 treatments at  $P < 0.05$ .

641

642 **Figure 7. Multivariate regression tree analysis of the dominant genera abundance and**  
643 **soil physicochemical variables.** Treatments and the number of samples included in the  
644 analysis are shown at the bottom. Abbreviations: AK, available potassium; AP, available  
645 phosphorus; C/N, carbon to nitrogen ratio; SOM, soil organic matter; TN, total nitrogen.

646

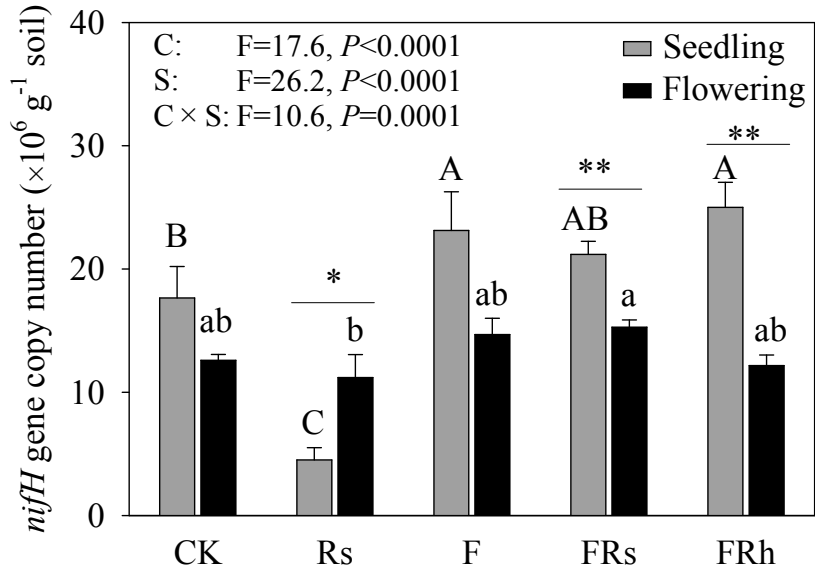
647 **Figure 8. Spearman correlation analysis between the relative abundances of dominant**  
648 **diazotrophic genera and soil physicochemical variables at the seedling and flowering**  
649 **stages of Chinese milk vetch.** *r* indicates the correlation coefficient; \*,  $P < 0.05$ .

650 Abbreviations: AK, available potassium; AP, available phosphorus; C/N, carbon to nitrogen  
651 ratio; SOM, soil organic matter; TN, total nitrogen.

652

653





**Figure 1**

655

656

657

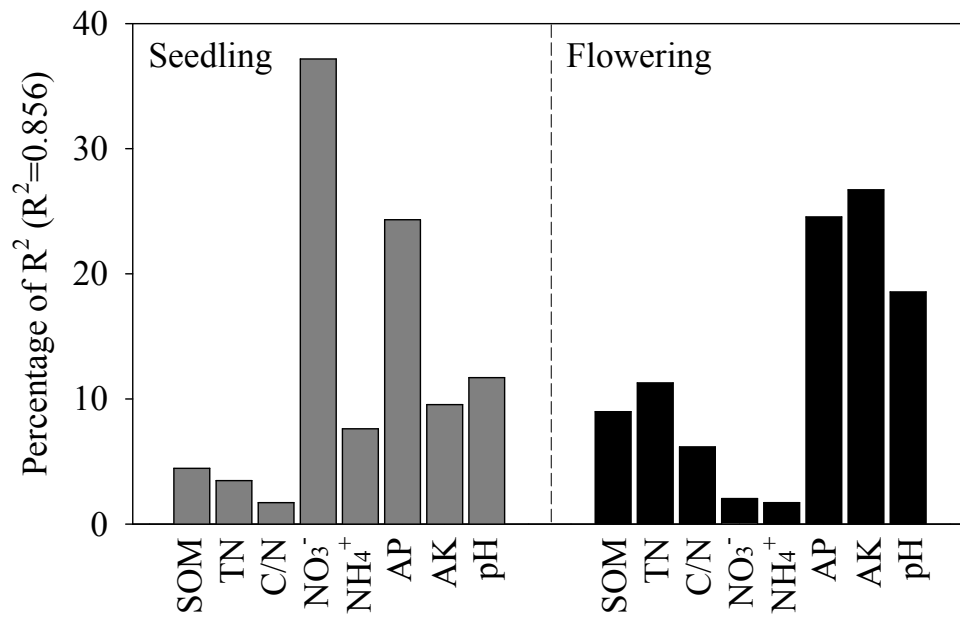
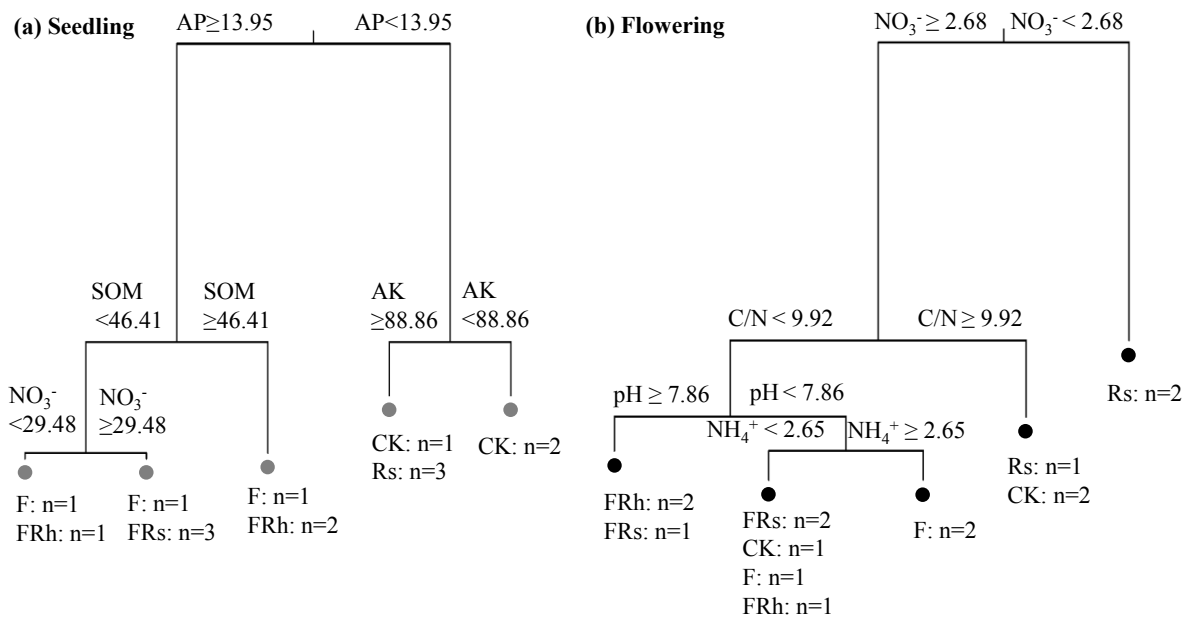


Figure 2

658

659

660

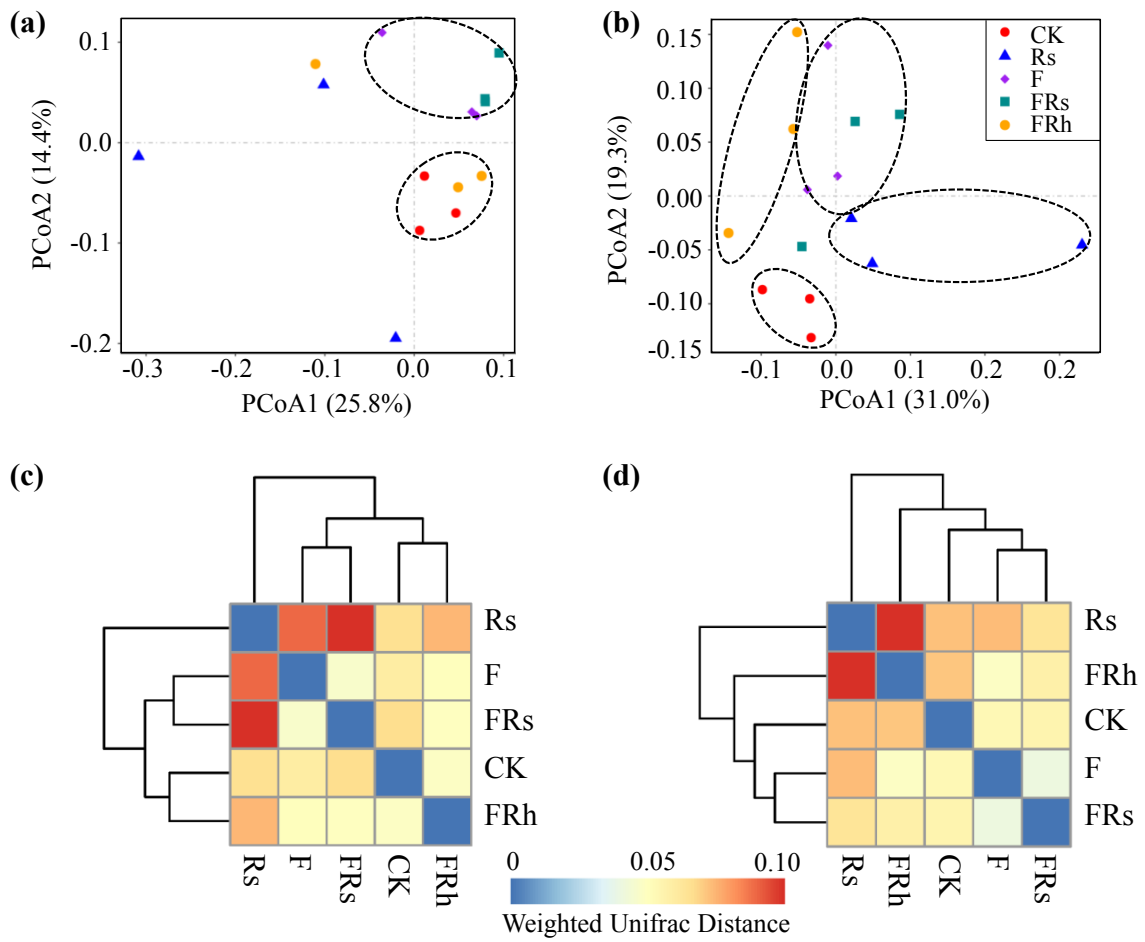


**Figure 3**

661

662

663



664

665

666

**Figure 4**

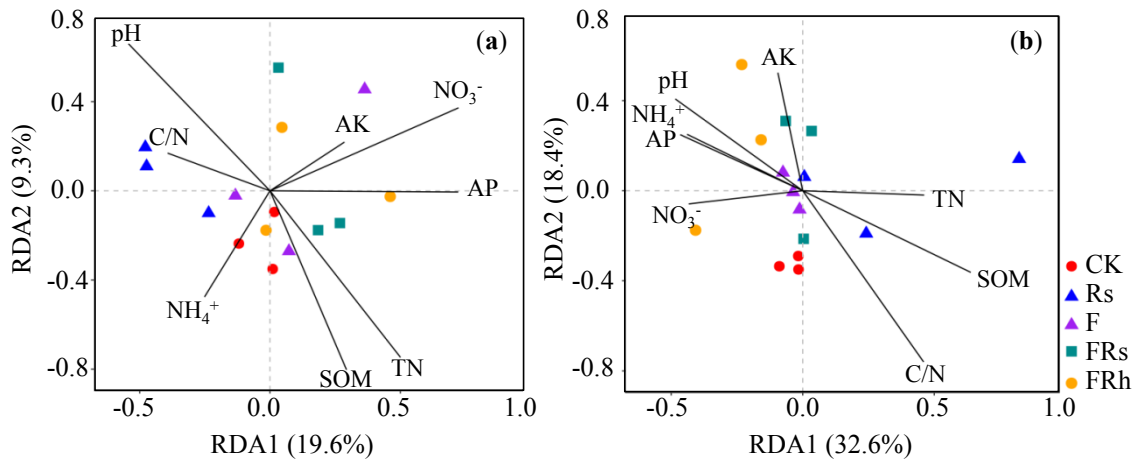


Figure 5

667

668

669

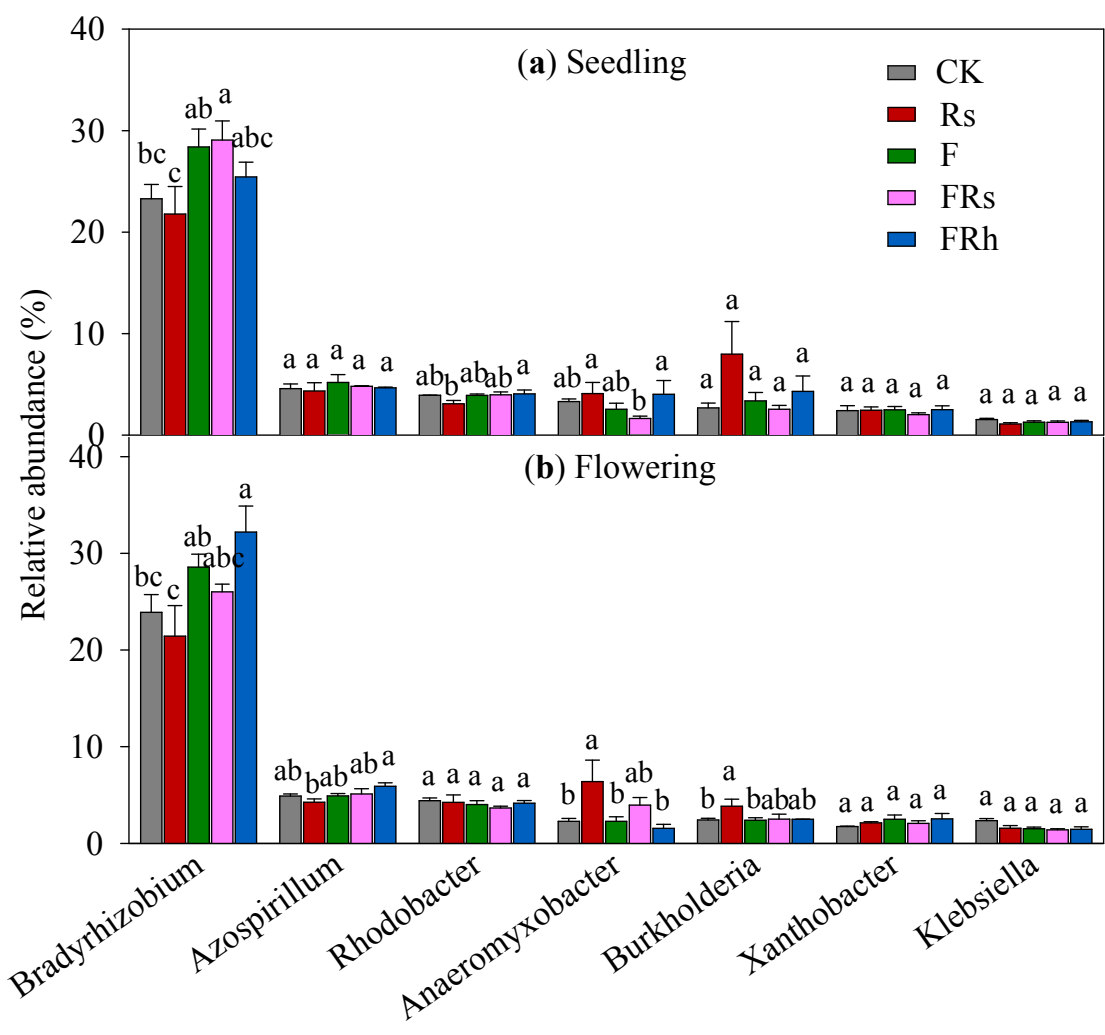
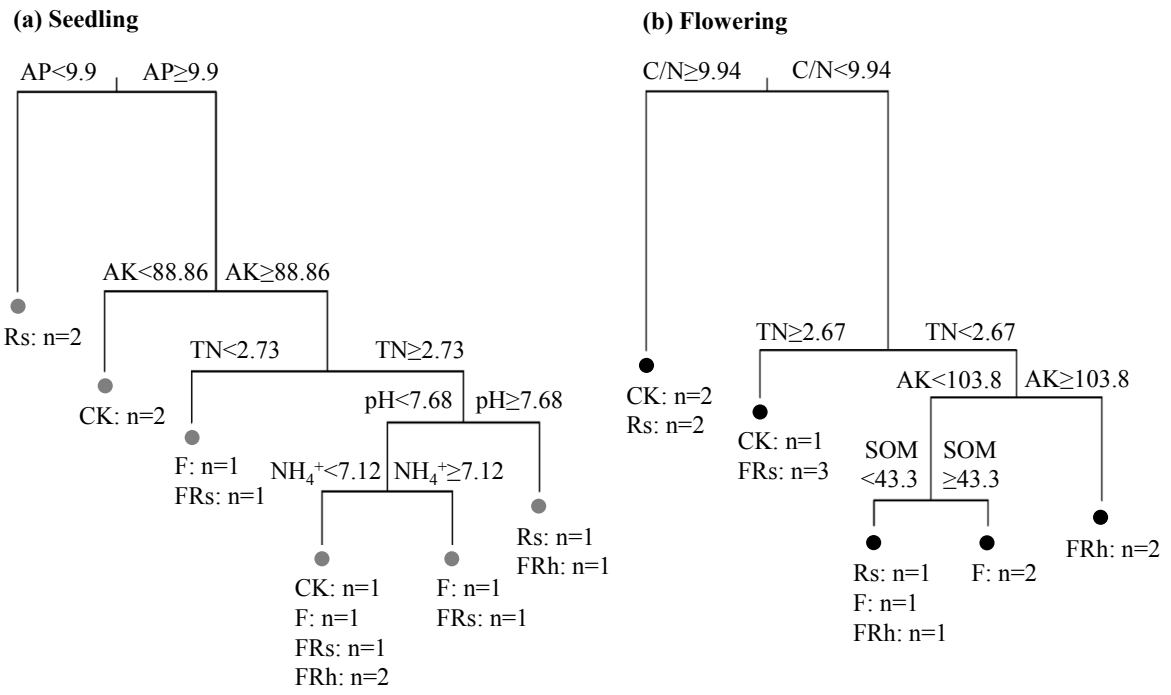


Figure 6

670  
671  
672

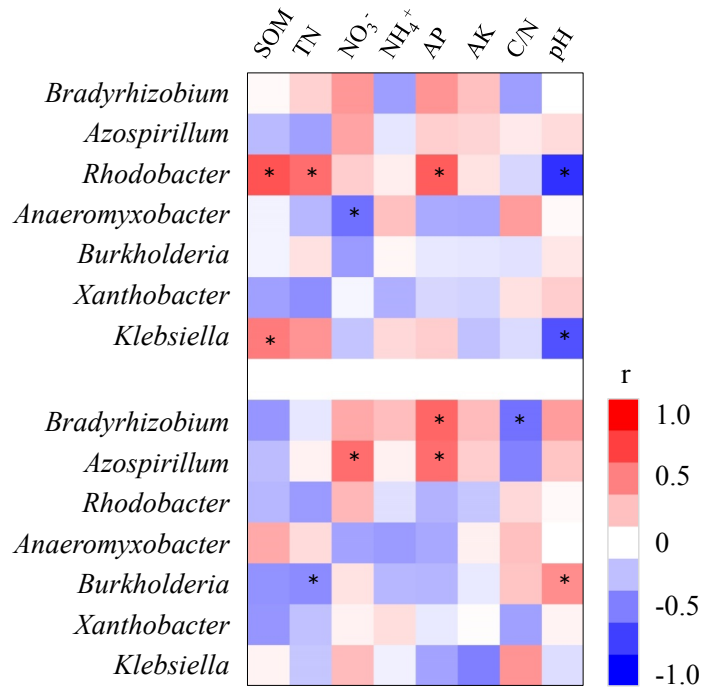


**Figure 7**

673

674

675



**Figure 8**

676

677

678



679 **Appendix A.** Supplementary data

680

681 **Table A1**

682 The estimated amount (kg ha<sup>-1</sup>) of exogenous N, P and K inputs by chemical fertilizer  
 683 application (N<sub>CF</sub>, P<sub>CF</sub>, and K<sub>CF</sub>) and biological N fixation of Chinese milk vetch (N<sub>BF</sub>) each  
 684 year #.

Treatment	Mv season	Early-rice season			Late-rice season			Annually total input		
	N <sub>BF</sub>	N <sub>CF</sub>	P <sub>CF</sub>	K <sub>CF</sub>	N <sub>CF</sub>	P <sub>CF</sub>	K <sub>CF</sub>	N	P	K
CK	45	0	0	0	0	0	0	45	0	0
Rs	47	0	0	0	0	0	0	47	0	0
F	47	120	26	60	144	16	80	311	42	140
FRs	46	120	26	60	144	16	80	310	42	140
FRh	51	120	26	60	144	16	80	315	42	140

685 # The nutrient input by precipitation and irrigation was not included, and the NPK brought by  
 686 straw return was considered as nutrient cycling within the double rice rotation system, not an  
 687 exogenous input.

688

689 **Table A2**

690 The two-way analysis of variation for the soil properties influenced by stage and cultivation  
 691 practice in this Chinese milk vetch-based system

Soil properties	Stage (S)		Cultivation practice (C)		S × C	
	F	<i>P</i>	F	<i>P</i>	F	<i>P</i>
Soil organic matter	0.16	0.691	0.24	0.9118	1.37	0.285
Total N	1.61	0.220	1.01	0.427	0.87	0.499
NO <sub>3</sub> <sup>-</sup>	375.75	<b>&lt;.0001</b>	2.44	0.0846	2.41	0.0868
NH <sub>4</sub> <sup>+</sup>	16.72	<b>&lt;.0001</b>	0.15	0.959	0.60	0.668
Available P	116.11	<b>&lt;.0001</b>	30.16	<b>&lt;.0001</b>	0.89	0.491
Available K	61.18	<b>&lt;.0001</b>	56.24	<b>&lt;.0001</b>	2.92	0.050
C/N ratio	3.77	0.068	3.61	<b>0.025</b>	1.52	0.237
pH	42.63	<b>&lt;.0001</b>	2.29	0.099	2.67	0.066

692

693