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8 3 Multi-day water residence time as a mechanism for physical and biological gradients  
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10 4 across intertidal flats  
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58  
59 **19 Abstract**  
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61 20 Tidal flats with shallow-sloping bathymetry under meso- to macrotidal conditions allow  
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63 21 organisms to occupy similar tidal elevations at different distances from subtidal  
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65 22 channels. As water floods or ebbs across such tidal flats during a single tidal cycle,  
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67 23 upstream organisms may modify water properties such as chlorophyll concentration,  
68  
69 24 while physiochemical properties may change due to close association with sediments.  
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71 25 Here we report evidence for an additional mechanism establishing cross-shore gradients:  
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73 26 multi-day water residence times, in the sense that even if water completely drains into  
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75 27 subtidal channels at low tide, a large fraction returns to the flats on the next high tide.  
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77 28 We applied circulation modeling and empirical measurements of water properties and  
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79 29 benthic secondary production to a 1-km-wide tidal flat in Willapa Bay, Washington,  
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81 30 USA. From the circulation model, water parcels on this intertidal flat have residence  
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83 31 times up to 2 d, that is, water found on the flat at one high tide returns to the intertidal  
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85 32 zone for a median of 4 successive semidiurnal high tides. Modeled residence times  
86  
87 33 generally increased towards shore. Four empirical datasets showed cross-shore gradients  
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89 34 consistent with modeled residence times: Salinity time series lagged towards shore;  
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91 35 water column chlorophyll declined towards shore at fixed stations (near-bottom) and in  
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93 36 surface transects more than could be explained by benthic suspension-feeding during a  
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95 37 single transit of water; and oyster (*Magallana = Crassostrea gigas*) condition declined  
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97 38 25% over 0.5 km from channel to shore, independent of tidal elevation. One  
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99 39 environmental measurement was more consistent with within-tide change, as water  
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101 40 temperatures warmed towards shore on afternoon flood tides but showed no tidal-cycle  
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103 41 lags. Taken together, these patterns suggest that multi-day water residence times can  
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115 42 contribute to environmental heterogeneity from channel to shore on tidal flats, acting  
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117 43 orthogonally to well-recognized estuarine gradients in residence time from ocean to  
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119 44 river.  
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123  
124 46 Keywords: benthic suspension feeders; circulation model; *Crassostrea gigas*; intertidal  
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126 47 gradients; residence time; water column chlorophyll  
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## 129 130 49 **1 Introduction**

131  
132 50 Coastal-plain estuaries and tidal embayments typically show systematic variation in  
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134 51 residence time and water age along the main axis from ocean mouth to head. The along-  
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136 52 channel residence-time gradient, which summarizes the net effect of various circulation  
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138 53 and mixing processes over a number of tidal cycles, broadly impacts biological and  
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140 54 biogeochemical estuarine dynamics. Increased residence time increases the fraction of  
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142 55 nitrogen that is denitrified (Dettmann, 2001), modifies sediment grain size (Wiberg et  
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144 56 al., 2015), reduces larval dispersal (Abelson and Denny, 1997), and results in reductions  
145  
146 57 of water column chlorophyll (Alpine and Cloern, 1992; Dame and Prins, 1998; Banas et  
147  
148 58 al., 2007). In estuaries with broad intertidal areas, it is much less common to analyze  
149  
150 59 cross-shore gradients (from the main channel to shore across a tidal flat) in terms of  
151  
152 60 residence time, as opposed to other schemas like tidal elevation or wave exposure. One  
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154 61 might well assume, in fact, that the residence time of the intertidal zone is, by definition,  
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156 62 at most a few hours, between one flood tide and the next ebb, and therefore simply not  
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158 63 commensurate with the multi-day residence times commonly seen on larger scales and  
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160 64 in deeper water. This study combines observations in Willapa Bay, Washington, USA  
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171 65 with semi-idealized numerical modeling to present a counterexample, in which  
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173 66 residence-time gradients provide a key mechanism for gradients in water properties and  
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176 67 secondary production across an intertidal mudflat.

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178 68 An earlier model of Willapa Bay (Banas and Hickey, 2005; Banas et al., 2007)  
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180 69 predicted strong residence-time gradients orthogonal to the estuarine axis, but those  
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182 70 studies did not have a means to validate that finding, or explore its biological  
183  
184 71 implications. These gradients did not reflect zonation associated with tidal elevation, but  
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186 72 rather a circulation pattern in which a large fraction of the water that ebbs off a flat into  
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188 73 the main channel returns on the subsequent flood tide, and in which the fraction returned  
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190 74 is greater for water found close to shore at high tide.

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192 75 Such small-scale variation in residence time may influence the productivity of  
193  
194 76 benthic organisms that depend on delivery of water column resources. Energy budgets  
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196 77 for Pacific oysters (*Magallana = Crassostrea gigas* Thunberg), as well as statistical  
197  
198 78 models relating oyster growth to environmental conditions, reveal strong effects of food  
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200 79 resources (Ren and Ross, 2001; Gangnery et al., 2003), water flow (Lenihan et al.,  
201  
202 80 1996), temperature, and salinity (Brown and Hartwick, 1988; Whyte et al., 1990; Ruiz et  
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204 81 al., 1992). Food quantity changes dynamically as particle concentrations are reduced  
205  
206 82 through grazing or increased through cell division or resuspension of benthic particles.  
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208 83 At small scales, individual performance may thus decline as density of benthic  
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210 84 suspension-feeders increases (Peterson and Black, 1987), and at larger scales those  
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212 85 individuals that are downstream may experience lower particle concentrations because  
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214 86 upstream individuals have already removed some (Grizzle et al., 2008).  
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227 87 Intertidal zonation in soft sediments is well established (Peterson, 1991; Dittman,  
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229 88 2000; Ryu et al., 2011), as is the effect of immersion time on performance within species  
230  
231 89 (e.g. bivalves Ruesink et al., 2003; Bishop and Peterson, 2006; Tomiyama et al., 2010;  
232  
233 90 Walles et al., 2016; Lomovasky et al., 2018). Yet in addition to the manifest  
234  
235 91 ramifications of how long water covers a particular intertidal point, the properties of that  
236  
237 92 water also shape the environmental context experienced by organisms. Thus an  
238  
239 93 understanding of the circulation and retention of water on tidal flats, which may underlie  
240  
241 94 heterogeneous water properties, becomes essential. An important distinction is  
242  
243 95 illustrated conceptually in Fig. 1. For water crossing a tidal flat during a single incoming  
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245 96 tide, water column resources may be filtered out during passage across beds of  
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247 97 suspension-feeders, resulting in downstream individuals with lower resource availability  
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249 98 (Fig. 1a). However, from a tidally-averaged perspective, some parcels of water may be  
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251 99 influenced by benthic suspension-feeders over multiple tides, and those portions of the  
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253 100 tidal flat with longer residence times may consequently have depleted water column  
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255 101 resources (Fig. 1b).

259 102 In addition to enhancing gradients in some water properties through longer  
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261 103 interactions with the benthos, water residence time on tidal flats has the potential to  
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263 104 generate lags in conservative tracers. During summer conditions of low riverflow,  
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265 105 especially at the mouth of Willapa Bay, salinity varies primarily due to the source water  
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267 106 that is tidally advected from the ocean, with salinity rising during upwelling, and falling  
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269 107 during downwelling (Roegner et al., 2002; Hickey et al., 2002; Ruesink et al., 2015).  
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271 108 Overall, water that has experienced an extended residence time could therefore be higher  
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273 109 or lower in salinity than “newer” water, depending on its origins during upwelling or  
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283 110 downwelling conditions. High residence time should consistently lead to warmer water  
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285 111 in summer, due to solar heating of dark tidal flats (Harrison and Phizacklea, 1987;  
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287 112 Hickey and Banas, 2003).

289  
290 113 In this paper we contribute evidence that the retention of water over multiple  
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292 114 tidal cycles, combined with the feeding activity of suspension-feeders, causes food  
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294 115 limitation in the intertidal zone, thus affecting secondary production. Oyster growers  
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296 116 respond to spatial variation in oyster growth at our study site by moving oysters in the  
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298 117 intertidal zone from shore to channel for fattening (improved meat weight; Hedgpeth  
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300 118 and Obrebski, 1981). Our focal questions were:

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302 119 1) What is the pattern of water residence time across this intertidal flat based on  
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304 120 circulation modeling?  
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306 121 2) Are channel-to-shore gradients in water properties (salinity, temperature, chlorophyll)  
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308 122 consistent with an extended water residence time?  
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310 123 3) How variable are oyster growth and condition from channel to shore, controlling for  
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312 124 tidal elevation?

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315 125 The overall goal is therefore to evaluate a previously unexplored mechanism of intertidal  
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317 126 water residence time in establishing cross-shore physical and biological gradients on  
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319 127 tidal flats.

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## 322 323 129 **2 Methods**

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### 326 327 328 131 *2.1 Study site*

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339 132 Willapa Bay, Washington, USA, has extensive tidal flats, with half of the bay area out of  
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341 133 the water on extreme low tides. We selected a tidal flat to study near the bay mouth  
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343 134 (46.59N, 124.02W, Fig. 2, 3), where much of the flat is occupied by commercial on-  
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345 135 bottom oyster culture, supported by plankton blooms advected from the nearshore ocean  
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347 136 (Roegner et al., 2002). Commercial shellfish aquaculture, primarily for Pacific oysters,  
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349 137 occupies approximately 20% of Willapa Bay's intertidal area (Feldman et al., 2000),  
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351 138 yielding up to 17% of the oysters cultured in the United States (Dumbauld and McCoy,  
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353 139 2015). Within ca. 50 ha at our study site, 38,000 bushels of oysters are harvested  
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355 140 annually (F. Wiegardt, pers. comm.).  
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## 360 142 *2.2 Circulation model of residence time*

361  
362 143 Banas and Hickey (2005) presented and validated a 255-m-resolution circulation model  
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364 144 of Willapa Bay, run under a variety of tide, riverflow, and wind forcing conditions.  
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366 145 More recently, a preliminary coupled bio-physio-chemical model of Willapa Bay,  
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368 146 implemented in ROMS (Regional Ocean Modeling System: Haidvogel et al., 2008) at  
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370 147 500 m resolution, has been introduced as part of the LiveOcean system  
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372 148 (<https://faculty.washington.edu/pmac/LO/LiveOcean.html>) and is being used to  
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374 149 produce daily forecasts. The model used in this study is a branch of the Banas and  
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376 150 Hickey (2005) model, implemented in ROMS but independent of the LiveOcean project,  
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378 151 and designed not for realistic hindcasting but for process insight, in the same spirit as a  
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380 152 tabletop fluid-dynamical lab experiment. Compared with the original Banas and Hickey  
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382 153 (2005) model, the ROMS model used here has simplified external forcing but, crucially,  
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384 154 much higher spatial resolution (50 m) and updated intertidal bathymetry.  
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395 155 The base bathymetry used in the model is the same as that used by Banas and  
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397 156 Hickey (2005), a dataset provided by the US Army Corps of Engineers Seattle District,  
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399 157 based on a survey of Willapa Bay's subtidal channels in 1998 (Kraus, 2000). Into this  
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401 158 model grid we substituted an improved intertidal bathymetric dataset, provided by the  
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403 159 Olympic Natural Resources Center (ONRC), which merges NOAA Coastal Service  
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405 160 Center LiDAR with locally collected point soundings and vertical datum transformations  
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407 161 for integration with USGS National Hydrographic Datasets by the ONRC staff. The final  
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409 162 model bathymetry uses the ONRC bathymetry at depths between 1.55 m above and 1.55  
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411 163 m below mean sea level (1.55 is the average of the difference between mean tide level  
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413 164 and mean lower low water (MLLW) at 7 NOAA sites around the bay), and the Kraus  
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415 165 (2000) model grid at deeper depths.  
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419 166 The model case used is a semi-idealized representation of summer, low-riverflow  
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421 167 conditions, in which the circulation is forced only by the semidiurnal (M2) tide and its  
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423 168 interaction with complex bathymetry, including wetting and drying of intertidal banks  
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425 169 (Oey, 2005; Warner, 2010; Xue and Du, 2010). For efficiency and stability, this  
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427 170 implementation of the model (unlike the original) is two-dimensional, i.e., barotropic:  
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429 171 this allows us to resolve fine-scale bathymetry with less smoothing, although some flow  
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431 172 information is lost. Banas et al. (2004) showed that neglecting baroclinic processes was  
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433 173 a fair approximation for late-summer, low-riverflow conditions in Willapa Bay, although  
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435 174 this simplification would not be appropriate for winter or spring conditions. M2 tidal  
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437 175 amplitude at the open boundary was set at 1.2 m. This produces a standard deviation  
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439 176 (SD) in sea level of 0.85 m at Toke Point near the bay mouth (NOAA station 9440910,  
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451 177 46° 42.5' N, 123° 58' W), which matches the observed SD of sea level at Toke Point  
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453 178 over three years to within 1%.

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455 179 Net circulation, residence time, and horizontal tidal diffusivity (a measure of the  
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457 180 strength of tidal stirring and the residual, tidally-averaged circulation) were calculated  
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459 181 based on the statistics of 170,000 particles (one per 50 m x 50 m grid cell) tracked for  
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461 182 one tidal cycle using depth-averaged currents. Beginning and ending positions of the  
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463 183 particles were used to construct a transition matrix or “tidal return map” (Banas and  
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465 184 Hickey, 2005; Banas et al., 2009) from which longer trajectories and residence-time  
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467 185 statistics were calculated. Residence times are reported below at 200 m resolution, based  
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469 186 on clusters of 16 particles released with 50 m spacing. Residence time is here defined as  
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471 187 the length of time that more than half of the 16 particles released in each 200 m square at  
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473 188 high tide continue to be found in the intertidal zone at successive high tides (where each  
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475 189 successive high tide in this calculation represents one set of lookups in the 50 m-  
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477 190 resolution return map).

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479 191 The model was validated by two methods: first, point comparisons with velocity  
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481 192 time series in the main channel and from the intertidal study site; and second, for a more  
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483 193 integrative measure, comparing tidal-excursion-scale horizontal diffusivity in the main  
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485 194 channel with empirical values based on salinity time-series analysis (Banas et al., 2004).  
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487 195 For the velocity validation, six velocity time-series stations were used, four in the main  
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489 196 channel as described by Kraus (2000) and previously used for model validation by  
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491 197 Banas and Hickey (2005), and two new intertidal stations, one towards shore  
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493 198 (N46.59774°, W124.03021°) and one near the channel (N46.59790°, W124.02082°).  
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495 199 Currents at these two intertidal stations were measured with acoustic Doppler current  
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507 200 profilers (Nortek Aquadopp) on 19 to 28 Jul 2008. Flow was determined for multiple 5  
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509 201 cm bins (north/south, east/west and up/down) beginning 10 cm off the bottom to within  
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511 202 10 cm of the surface of the water, and measurements were taken at 2 MHz at 10 minute  
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513 203 intervals with a 0.05 m blanking distance. At both stations, water depth at mean high  
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515 204 water was 2 m. Depth-averaged root mean square (rms) tidal velocity at all six stations  
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517 205 from observations and the model are given in Table 1. Percent errors range from 1-35%  
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519 206 with a mean of 15%, generally increasing up-estuary and shoreward, and generally in the  
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521 207 direction of overestimated velocities and underestimated velocity gradients in the model.  
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523 208 Note that this mode of error probably biases model results in a direction opposite to our  
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525 209 conclusion that strong net-circulation gradients exist across intertidal flats.  
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529 210 For the horizontal diffusivity analysis, which measures the net, tidally averaged  
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531 211 tidal circulation (as opposed to the amplitude of tidal currents themselves), we calculated  
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533 212 diffusivities  $K_H$  at the five main-channel stations where observational estimates were  
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535 213 previously reported (Banas et al., 2004), from the rate of horizontal dispersion of square  
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537 214 patches of model particles the same width as the channel:  $K_H = 1/2 d\langle x^2 \rangle / dt$ , where  $t$  is  
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539 215 time and  $\langle x^2 \rangle$  is the two-dimensional variance in particle positions around their center  
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541 216 of mass. Like the Banas and Hickey (2005) model, this new version of the model  
542  
543 217 replicates the high diffusivities (200-700  $\text{m}^2 \text{s}^{-1}$ ) observed in moored salinity time series  
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545 218 in the well-flushed outer 20 km of the estuary, where our study site is located.  
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548 219 In the poorly flushed, southern reaches of the bay, where observationally-derived  
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550 220 horizontal diffusivities on the scale of the channel width are 50-100  $\text{m}^2 \text{s}^{-1}$  (Banas and  
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552 221 Hickey, 2005), the new ROMS model underestimates diffusivities by approximately a  
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554 222 factor of two and predicts main-channel residence times ~25-50% higher than the Banas  
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563 223 and Hickey (2005) estimate. These discrepancies could result from both models' simple  
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565 224 assumptions about bottom friction; bias in the depth of shallow tidal flats introduced in  
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567 225 the merging of intertidal and subtidal bathymetric surveys; under-resolution of narrow,  
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569 226 secondary and tertiary channels; or weak baroclinic effects. Since these issues do not  
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571 227 seem to affect our study area and in any case would be difficult to pursue without  
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573 228 extensive new observations, we have simply confined our analysis to the middle-to-outer  
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575 229 estuary (Fig. 2), and refrain from speculating about residence time patterns across tidal  
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577 230 flats in the southern bay.  
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### 582 232 *2.3 Cross-shore pattern of temperature and salinity*

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585 233 Salinity can act as a passive tracer of water age, particularly when source water varies in  
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587 234 salinity due, in this case, to coastal upwelling strength. We deployed multiparameter  
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589 235 dataloggers (YSI Datasonde 6600) between 2 Jun and 28 Aug 2008 to record salinity (as  
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591 236 specific conductivity), temperature, and water depth at five stations spanning 1 km from  
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593 237 channel to shore (Fig. 3a, Table S1). Probes were suspended 0.1 m above the sediment,  
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595 238 recording at 10-min intervals, and were cleaned every 2 wk. Sensors were held in a  
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597 239 common water bath four times throughout the summer, and specific conductivity  
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599 240 adjusted for two of the sensors showing consistent offsets from the others. Specific  
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601 241 conductivity (mS/cm) was then converted to salinity (practical salinity units; Wagner et  
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603 242 al., 2006). For the three-month period of deployment, median water levels were 1.6 m  
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605 243 relative to MLLW. Seven measurements centered around the time of median water level  
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607 244 were averaged on each flood tide at each sensor; this process standardized comparisons  
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610 245 across the tidal flat.  
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619 246 The Bakun upwelling index for the time series point closest to Willapa Bay (48°  
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621 247 N, 125° W) was used as an index of salinity in source water outside the bay  
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623 248 ([http://www.pfel.noaa.gov/products/PFEL/modeled/indices/upwelling/NA/data\\_downloa](http://www.pfel.noaa.gov/products/PFEL/modeled/indices/upwelling/NA/data_download.html)  
624  
625 249 [d.html](http://www.pfel.noaa.gov/products/PFEL/modeled/indices/upwelling/NA/data_download.html)). For each of two sensors closest to the channel, each day's salinity (mean of two  
626  
627 250 flood tides on most days) was related to the daily upwelling index using a linear model,  
628  
629 251 and model fit ( $r^2$ , P value for  $n=63$  or 68 days) was recorded for time lags of 0 to 7 days.  
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631 252 Due to summer drought, these models also included day of year as a predictor to account  
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633 253 for generally increasing salinity in the bay during the summer.  
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636 254 Two types of analyses were carried out to determine cross-shore gradients in  
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638 255 temperature and salinity, testing each of three shoreward stations against the channel  
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640 256 station with the most complete record (ChS). The first type emphasized mean  
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642 257 differences in water properties based on paired t-tests (paired by each flood tide). The  
643  
644 258 second type emphasized time lags in water properties by examining model fit ( $r^2$  for  
645  
646 259  $n\sim 75$ -130 flood tides) of the relationship between the times series at two sensors, with  
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648 260 lags from the channel sensor of 0 to 4 tidal cycles.  
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#### 652 262 *2.4 Cross-shore pattern of chlorophyll concentration*

653 263 Two approaches were taken to determine how chlorophyll concentration changed across  
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655 264 the tidal flat, one involving fixed sensors and a second by motoring a sensor along  
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657 265 transects from channel to shore. We deployed multiparameter dataloggers (YSI  
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659 266 Datasonde 6600) between 2 and 15 Aug 2007 to record chlorophyll fluorescence, water  
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661 267 depth, temperature, and salinity, at one near-channel station (ChN) and three towards  
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663 268 shore (two of these three sensors were at ShMid; Fig. 3a, Table S1). Sensors spanned  
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675 269 0.56 km, with probes suspended 0.1 m above the sediment and recording at 10-min  
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677 270 intervals. To make fluorescence readings consistent among sensors, the four sensors  
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679 271 were held in common conditions while chlorophyll was altered by adding and filtering  
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681 272 phytoplankton (14 levels), and adjustments were made to raw values so that all sensors  
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683 273 had the same slope and zero intercept; however, no bottle samples were collected for  
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685 274 calibration to chlorophyll-a, and so results are provided only as fluorescence. During  
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687 275 field deployment, the sensors occasionally generated segments of particularly noisy data,  
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689 276 in which both the values themselves and their change between 10-min intervals were  
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691 277 large (i.e. corrected values  $>25 \mu\text{g L}^{-1}$  and fluctuation  $>10 \mu\text{g L}^{-1}$ ), possibly from  
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693 278 catching drift macroalgae. We removed readings  $>25 \mu\text{g L}^{-1}$  before proceeding to the  
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695 279 next step of calculating tide-specific fluorescence. Seven measurements centered around  
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697 280 the time of each median water level (1.6 m MLLW) were averaged for each flood and  
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699 281 ebb tide for each sensor. We required at least three of the seven measurements to have  
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701 282 passed the  $25 \mu\text{g L}^{-1}$  filter, and then removed any values  $>10 \mu\text{g L}^{-1}$  from the mean  
702  
703 283 fluorescence of other measurements. Also, strong drift in fluorescence was evident at  
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705 284 one ShMid sensor for the last five days of deployment, and these means were included  
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707 285 in visual display of all data but not used in analysis. Once these steps to generate means  
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709 286 from reliable sections of fluorescence readings were complete, we calculated the  
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711 287 difference in fluorescence (t-test, paired by tidal cycle) between each sensor and the one  
712  
713 288 closest to the channel (ChN), separately for flood and ebb tides.

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715 289 Chlorophyll concentrations were mapped across the tidal flat an hour before the  
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717 290 afternoon high tide on 17 Aug 2008. This small boat-based sampling involved driving  
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719 291 six transects from channel to shore while water was forced into the opening of a tube  
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731 292 under the boat, 0.2 m below the surface. The water then fed into a pipe holding a YSI  
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733 293 Datasonde 6600, which measured temperature, salinity (as specific conductivity), and  
734  
735 294 fluorescence every 5 seconds. A GPS (Garmin Geko) simultaneously recorded position.  
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737  
738 295 To calibrate the chlorophyll sensor, bottle samples were collected (300 ml in triplicate at  
739  
740 296 three positions), extracted in 90% W/V acetone and frozen >24 h, and measured on a  
741  
742 297 fluorometer (Turner Designs AU-10) following acidification procedure (Welschmeyer  
743  
744 298 1994). This calibration showed that the fluorescence values recorded by the sensor  
745  
746 299 needed to be altered: Chlorophyll-a =  $0.319 \times \text{Fluorescence}$  ( $r^2=0.68$ ,  $N=9$ ). Each  
747  
748 300 transect of about 0.5 km had 14-31 measurements at different distances from the channel  
749  
750 301 (MLLW contour). Chlorophyll-a was considered a response variable and distance a  
751  
752 302 predictor variable in linear models (regression) to calculate the slope and standard error  
753  
754 303 for each transect. Then, meta-analysis procedures were applied to these six slopes and  
755  
756 304 SE to calculate the overall change in chlorophyll with distance from the channel (rma  
757  
758 305 command in package metafor; Viechtbauer, 2017).  
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### 763 307 *2.5 Cross-shore pattern of oyster performance*

764  
765 308 Growth and condition of juvenile oysters (*Magallana* = *Crassostrea gigas*) were  
766  
767 309 measured at five stations on the tidal flat where the overall bathymetry was sufficiently  
768  
769 310 flat to enable deployment at a common tidal elevation, judged by water level (Fig. 3a).  
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771 311 Distance to channel was determined for each station based on the MLLW contour, and  
772  
773 312 ranged from 0.2 to 0.7 km (Table S1). Hatchery-raised oyster larvae were settled onto 11  
774  
775 313 x 11 cm unglazed ceramic tiles, thinned to 8-15 oysters per tile, and grown to a size of 1  
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777 314 cm shell length in a common location. On 3 Aug 2007, five tiles were attached vertically  
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787 315 to PVC poles at an elevation of +0.6 m MLLW at each of the five outplant stations. Tiles  
788  
789 316 were always at least 0.15 m above the sediment, although this distance varied slightly  
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791 317 among stations. Tiles were collected 19 May 2008, and all oysters were measured for  
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793 318 maximum shell length from the umbo (shell height, mm). Subsequently, oyster meat was  
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795 319 removed and dried (60°C, nearest 0.01 g) and a metric of condition developed as the  
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797 320 ratio of dry meat weight to shell length. We were unable to remove bottom valves from  
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799 321 the tiles, thus precluding the use of typical condition index based on ratios of tissue mass  
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801 322 to internal shell volume (Lawrence and Scott, 1982). Of 25 tiles deployed, 23 were  
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803 323 recovered. Of 171 oysters measured, three were removed from analysis because their  
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805 324 tissue mass did not register on our balance (two from a shoreward station and one near  
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807 325 the channel) and an additional one was censored because its weight was probably  
808  
809 326 incorrectly recorded (order of magnitude more dry mass than any other oyster). Oyster  
810  
811 327 shell height and condition were analyzed with linear mixed effects models, in which  
812  
813 328 distance to channel was a fixed effect and tile was a random effect to account for  
814  
815 329 multiple oysters per tile (package nlme, Pinheiro et al., 2016). Statistical significance  
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817 330 was set at  $\alpha=0.05$ . Analyses of water properties and oyster performance were performed  
818  
819 331 in R (R Core Team, 2015). Empirical data underlying these analyses are archived at  
820  
821 332 <http://dx.doi.org/10.17632/wx9y9njnr.1>  
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### 828 334 **3 Results**

#### 829 335 830 336 *3.1 Water residence time*

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843 337 The ROMS numerical model showed a strong spatial pattern in intertidal residence time  
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845 338 around our study site. Fig. 2 depicts the number of tidal cycles it takes for half of the 16  
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847 339 particles originating in each 200 x 200 m square in the intertidal zone to be found  
848  
849 340 outside the intertidal zone at high water. This quantity is a proxy for the length of time  
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851 341 that intertidal grazers have access to a particular water parcel. In our study region, this  
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853 342 intertidal residence time varies from 0-4 tidal cycles over a distance of 2 km, with a  
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855 343 reduced range over the 1-km tidal flat where stations for empirical measurements were  
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857 344 located. Values are patchy, but generally increase toward shore. Gradients in residence  
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859 345 time emerge on a number of wide tidal flats in the middle-to-outer estuary (Fig. 2), not  
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861 346 only at our study site. However, the width of a particular tidal flat is not a reliable  
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863 347 predictor of the mean or maximum intertidal residence time: examples can be found  
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865 348 (Fig. 2) of narrow flats with residence time much greater than one cycle, and extensive  
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867 349 intertidal flats where the outer portions are well-flushed within one cycle, as one might  
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869 350 naively expect.  
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### 875 352 *3.2 Cross-shore pattern of temperature and salinity*

876  
877 353 Upwelled, high-salinity water outside the bay reached the channel sensors after a 4-day  
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879 354 lag (Fig. 4a, Table 2). The highest  $r^2$  values were associated with 4-day lags, and model  
880  
881 355 fit for salinity was significantly improved by including the Bakun upwelling index as a  
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883 356 predictor, lagged by four days (likelihood ratio tests of models with no upwelling  
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885 357 predictor vs. 4-day lag: ChN  $F_{1,60}=21.4$ ,  $P<0.0001$ , ChS  $F_{1,66}=21.3$ ,  $P<0.0001$ ).  
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887 358 Empirically, water properties at median flood tide differed between sensors deployed at  
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889 359 channel and shore stations (Fig. 4b, c). Water at the station closest to shore (ChIn) was  
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899 360 0.58°C warmer than at the channel, and other shoreward stations followed this pattern  
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901 361 (0.62°, 0.30°C warmer, Table 3). Salinity tended to be lower at shoreward stations  
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903 362 relative to the channel stations and also lagged by one tidal cycle from channel to shore  
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905 363 (Table 3). This lag is particularly evident in Fig. 4b for a portion of summer 2008 when  
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907 364 salinity was initially lower towards shore and then reversed to be higher towards shore,  
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909 365 as downwelled low-salinity water reached the shoreward portion of the tidal flat later  
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911 366 than it appeared near the channel. No lags were necessary in the best fit model for the  
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913 367 temperature time series, but a notable feature here was that lags of 0, 2, 4 tidal cycles fit  
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915 368 better than lags of 1, 3 cycles (Table 3). This evidently arose due to the diurnal cycle in  
916  
917 369 heating and cooling, with the warmer temperatures occurring during afternoon and early  
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919 370 evening.  
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### 925 372 *3.3 Cross-shore pattern of chlorophyll concentration*

926  
927 373 In general, fluorescence was low as water began to flood onto the tidal flat, then  
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929 374 increased during the flood and declined during the ebb, especially at the channel station  
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931 375 (Fig. 5c). On flood tides, the upstream station was near the channel and downstream  
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933 376 stations towards shore. All downstream stations had significantly lower fluorescence  
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935 377 than the channel station as water flooded through median water level (Fig. 5a, Table 4).  
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937 378 During flood tides, the sensor at ShOut recorded 78% of the fluorescence relative to the  
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939 379 channel station, and this proportion was even lower at the two ShMid sensors (37%,  
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941 380 64%). As the tide ebbed, the channel station was downstream from stations closer to  
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943 381 shore, but still had greater or similar fluorescence relative to shoreward stations (Fig. 5b,  
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955 Table 4). Fluorescence did not differ at ShOut and ChN, but one of the ShMid sensors  
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957 recorded lower fluorescence (47% relative to ChN).  
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960 Chlorophyll concentrations in surface water measured along transects also  
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962 declined from channel to shore (Fig. 3b, Table S2). Chlorophyll-a at the start of each  
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964 transect, closest to the channel, was  $4.23 \mu\text{g L}^{-1}$ , and this declined at  $-2.99 \mu\text{g L}^{-1} \text{ km}^{-1}$   
965  
966 (SE 0.53). Accordingly, over 0.5 km, the proportion of chlorophyll remaining was 65%.  
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### 968 969 970 3.4 Cross-shore pattern of oyster performance 971

972 Over nine months, juvenile oysters on tiles grew from about 1 cm to 3 cm. Final shell  
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974 height of juvenile oysters did not differ with distance to channel ( $t_{1,21}=-0.27$ ,  $p=0.8$ ; Fig.  
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976 6a). However, condition (dry tissue weight per shell height) followed a channel-to-shore  
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978 gradient (Fig. 6b). Condition declined 25% across the intertidal flat as oysters were 0.5  
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980 km further from the channel (Condition =  $5.73 - 2.828 \times \text{Distance in km}$ ;  $t_{1,21}=-2.66$ ,  
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982  $p=0.015$ ). This pattern emerged even though tidal elevation and immersion time were  
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984 held constant.  
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## 987 988 989 4 Discussion 990

991 The key result to emerge from both modeling and empirical work in our study is that  
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993 heterogeneous water residence time on intertidal flats provides an additional mechanism  
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995 for cross-shore physical and biological gradients, beyond the more obvious mechanisms  
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997 of tidal elevation and upstream-downstream position, and despite water draining off the  
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999 flat at each low tide. At the subtidal (tidally-averaged) scale, water found near shore at  
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1001 high tide tends to return to the intertidal zone every high water for a median of 4 tidal  
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1011 405 cycles, whereas water near the channel is largely replaced every tidal cycle (Fig. 2). By  
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1013 406 examination of other tidal flats modeled in middle-to-outer Willapa Bay, it is clear that  
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1015 407 intertidal residence time is not a simple function of distance to channel or bathymetry;  
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1017 408 however, gradients in residence time only developed on wider ( $\geq 1$  km) tidal flats, while  
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1019 409 heterogeneity appeared in model results at smaller scales (Fig. 2).  
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1022 410 Residence-time gradients across tidal flats have two implications. First, food  
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1024 411 depletion must be thought of as a several-tidal-cycle process: although the water found  
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1026 412 on the tidal flat drains into the channel on every ebb tide, its return over several  
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1028 413 successive flood tides allows it to be repeatedly and intensively modified through  
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1030 414 interactions with the sediment and benthic organisms. Second, because of this process,  
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1032 415 tidal circulation may create gradients in food availability over an intertidal area of 1 km  
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1034 416 or less, because differences in water age allow neighboring water parcels to be subject to  
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1036 417 different levels of depletion even if grazing pressure is spatially uniform.  
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1039 418 Do such multi-day water residence times, as predicted by the numerical model  
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1041 419 reported here, as well as a prior iteration (Banas et al., 2007), actually occur on intertidal  
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1043 420 flats? Observations of physical (salinity) and biological (chlorophyll) water properties  
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1045 421 indicate that they do. Weather-event-scale fluctuations in salinity show a time lag (one  
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1047 422 tidal cycle) from channel to shore, indicating that the water towards shore takes longer to  
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1049 423 be replaced than water near the channel: the tidal flat is not renewed as a single unit on  
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1051 424 each flood tide. Note that the lag of 1 tidal cycle was measured over a smaller distance  
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1053 425 than the full tidal flat width over which the model predicts a residence time gradient  
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1055 426 from 0–4 cycles, and also that the lag time is a different physical quantity (signal  
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1067 427 propagation time through an oscillatory boundary layer: Batchelor, 1967) from a  
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1069 428 residence time estimate.

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1071 429 For chlorophyll, it is necessary to ask whether the observed gradients could have  
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1073 430 been achieved through suspension-feeding as the water transited between sensors during  
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1075 431 a single flood tide (as in Fig. 1a), or whether a longer period of interaction is required  
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1077 432 (Fig. 1b). Key parameters to distinguish these processes are available from prior studies  
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1079 433 of this tidal flat in which parcels of water were tracked with Lagrangian drifters (Wheat  
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1081 434 and Ruesink, 2013): exponential loss rates of chlorophyll were  $-0.24 \text{ h}^{-1}$  due to the  
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1083 435 feeding of benthic suspension-feeders, while water velocity averaged  $0.16 \text{ m s}^{-1}$ ,  
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1085 436 consistent with model results and stationary sampling (Table 1). The loss rate of  $-0.24 \text{ h}^{-1}$   
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1087 437 was measured at water depths  $<1.2 \text{ m}$  (Wheat and Ruesink, 2013), and so represents an  
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1089 438 upper bound (i.e. rapid decline) in which suspension feeders affect a small volume of  
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1091 439 overlying water. Recall that ShMid sensors showed 37% and 64% of fluorescence  
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1093 440 relative to ChN, and these sensors were separated by  $0.56 \text{ km}$  east-west ( $1 \text{ h}$  at  $0.16 \text{ m s}^{-1}$ ),  
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1095 441 giving loss rates of  $-0.99$  and  $-0.44 \text{ h}^{-1}$ . Thus the observed decline was two to four  
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1097 442 times greater than could be explained by a single passage of water between the stations.  
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1099 443 Additionally, during ebb tides, stations near the channel were in a downstream position  
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1101 444 relative to suspension-feeding occurring on the tidal flat. However, compression, rather  
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1103 445 than reversal, of the channel-to-shore gradient occurred on ebb relative to flood tides  
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1105 446 (Fig. 4). That is, on ebb tides, the shoreward stations were still lower in fluorescence  
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1107 447 than expected from a simple consideration of water influenced during a single transit  
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1109 448 between stations.  
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1123 449 Transects from channel to shore showed that chlorophyll declined to 65% over  
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1125 450 0.5 km (0.87 h at 0.16 m s<sup>-1</sup>; Fig. 3b, Table S2), a loss rate of -0.50 h<sup>-1</sup>. As such,  
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1127 451 empirical data from transects concurred with fixed sensors in demonstrating greater  
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1129 452 chlorophyll losses in water towards shore than expected from suspension-feeding in  
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1131 453 even-aged water. The results were consistent despite different methods, e.g.,  
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1133 454 measurements on transects within 0.2 m of the water surface, but within 0.1 m of the  
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1135 455 sediment for fixed sensors. Fixed sensors gave further insight into mechanisms  
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1137 456 underlying gradients in chlorophyll concentrations through examination of the time  
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1139 457 series at 10-min resolution. Specifically, the first water to flood onto the tidal flat and  
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1141 458 reach the stations near the channel was low in fluorescence (Fig. 5c), suggesting older,  
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1143 459 depleted water that mixed only weakly with newer, chlorophyll-rich water while in the  
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1145 460 channel during low slack tide (cf. MacDonald, 2006).

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1147  
1148 461 Our conclusion that cross-shore gradients in chlorophyll were established  
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1150 462 through multiple passes of water over the tidal flat needs to be evaluated in light of  
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1152 463 several other factors known to limit draw-down by suspension feeders. Growth  
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1154 464 dynamics of phytoplankton (Calbet and Landry, 2004) and/or resuspension from the  
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1156 465 benthos (Ruesink et al., 2019) would tend to make our loss rates underestimates. Effects  
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1158 466 of suspension feeders on chlorophyll are also expected to be diluted as water depth  
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1160 467 increases. Overall, each of these factors would make residence-time variation less  
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1162 468 important to cross-shore gradients in chlorophyll, yet we nevertheless found empirical  
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1164 469 gradients greater than could be explained by benthic suspension-feeding during a single  
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1166 470 transit of water.  
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1179 471 In contrast to salinity and chlorophyll, which supported that water appeared  
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1181 472 multiple times on the tidal flat, temperature was modified primarily within tidal cycles.  
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1183 473 We draw this conclusion due to the absence of statistical lags between channel and  
1184  
1185 474 shoreward stations (Table 3), but also from strong diel variability in temperature. Water  
1186  
1187 475 temperatures were warmer during daytime than nighttime flood tides, and also exhibited  
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1189 476 stronger channel-to-shore gradients during the day (Fig. 4c). Indeed, the first water to  
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1191 477 reach the shoreward sensor as the tide flooded in late afternoon was overall highest in  
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1193 478 temperature (Fig. 5d), as its small volume was heated by contact with dark sediment. In  
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1195 479 aggregate, these observations are consistent with water being heated as it travels over  
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1197 480 solar-heated sediment and cooled at night, making residence time effects less apparent.  
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1199 481 The distinction may arise because salinity is a conservative tracer and chlorophyll is  
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1201 482 likely to accumulate benthic influence in one direction only, whereas water temperatures  
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1203 483 may be modified in both positive and negative directions by a variety of heat-budget  
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1205 484 drivers acting at shorter time scales than residence time.

1206  
1207 485 Food limitation due to draw-down of water column resources by other  
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1209 486 suspension feeders has been documented at two scales: at a scale of near neighbors, for  
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1211 487 instance in shellfish aquaculture stocked at high densities (Newell, 1990; Muschenheim  
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1213 488 and Newell, 1992; Senechal et al., 2008; Grangere et al., 2010), and at whole-basin  
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1215 489 scales when feeding by suspension feeders exceeds the delivery or growth of  
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1217 490 phytoplankton (Alpine and Cloern, 1992). In some cases, delivery of water column  
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1219 491 resources is not tightly linked to benthic secondary production, where the diet of  
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1221 492 suspension-feeders is dominated by resuspended microphytobenthos (Kang et al., 2003;  
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1223 493 Herman et al., 2000; van Oevelen et al., 2006). Nevertheless, competition impacts not  
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1235 494 only near neighbors, by creating local regions of food depletion (Lenihan, 1999;  
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1237 495 Grangere et al., 2010), but also the food availability for distant individuals when  
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1239 496 extended water residence time allows for significant filtration. Secondary productivity  
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1241 497 on the tidal flat is accordingly driven by a complex interaction between filtration and  
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1243 498 residence time dynamics. Other studies have found similarly complex relationships  
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1245 499 between plankton dynamics, water advection and benthic secondary productivity  
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1247 500 (Tweddle et al., 2005; Simpson et al., 2007). Despite these complexities, the connection  
1248  
1249 501 between (low) residence time and (high) bivalve carrying capacity has long been known  
1250  
1251 502 at whole-estuary scales (Dame and Prins, 1998; Zu Ermgassen et al., 2013). Lower  
1252  
1253 503 oyster condition is therefore consistent with longer water residence time towards shore,  
1254  
1255 504 controlling for tidal elevation. Although shell growth did not follow a cross-shore  
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1257 505 gradient (Fig. 6), condition is the primary economic benchmark on this tidal flat, which  
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1259 506 is a fattening ground for oysters (Hedgpeth and Obrebski, 1981).

1260  
1261  
1262 507 As expected in an estuary, salinity generally increased with water level as  
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1264 508 oceanic water was advected into the estuary on each flood tide (Fig. 5d). This is  
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1266 509 consistent with earlier conclusions (Roegner et al., 2002; Banas et al., 2007) that in this  
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1268 510 coastal-upwelling-driven system, phytoplankton are primarily supplied by the coastal  
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1270 511 ocean and progressively depleted within the estuary, such that high chlorophyll is  
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1272 512 correlated with high salinity, both across the tidal flat and across the tidal cycle within  
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1274 513 stations (Fig. 5c,d, Table S2).

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1276  
1277 514 Estuaries are typically described as heterogeneous in their along-axis dimension,  
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1279 515 including both physical and biological gradients (Attrill and Rundle, 2002; Ruesink et  
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1281 516 al., 2015; Tweedley et al., 2016). Orthogonal to this axis, as depth is reduced and flats  
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1291 517 emerge at low tide, water flow and residence time can be influenced by wind, small  
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1293 518 bathymetric features such as hummocks and sloughs, and seawater-porewater exchange  
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1295 519 (Pokavanich and Alosairi, 2014; Sullivan et al., 2015). To our knowledge, no prior  
1296  
1297 520 reports exist (other than Banas et al., 2007) of heterogeneity in multi-day residence time  
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1299 521 on intertidal flats, although extended residence times towards shore can be inferred from  
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1301 522 general principles if water reaching shallower depths is restricted in mixing with new  
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1303 523 ocean water (Hsu et al., 2013). Our empirical measurements of water properties support  
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1305 524 the existence of residence time variation at scales of 0.5-1 km intertidally, which  
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1307 525 effectively lengthens the time scale over which upstream suspension-feeders may  
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1309 526 modify resource delivery underlying benthic secondary production. The sedimentary  
1310  
1311 527 characteristics of tidal flats are now understood as a predictable function of sediment  
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1313 528 supply and ratio of tidal currents to waves (Gao, 2019); however, water properties across  
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1315 529 these flats require further scrutiny to determine which morphological and hydrodynamic  
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1317 530 factors may lead to multi-day residence times.  
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1349 541 **References**

1350  
1351 542 Abelson, A., Denny, M., 1997. Settlement of marine organisms in flow. *Ann. Rev. Ecol.*  
1352  
1353 *Syst.* 28, 317-339.  
1354 543

1355  
1356 544 Alpine, A.E., Cloern, J.E., 1992. Trophic interactions and direct physical effects control  
1357  
1358 phytoplankton biomass and production in an estuary. *Limnol. Oceanogr.* 37, 946–  
1359 955.  
1360 546

1361  
1362 547 Attrill, M.J., Rundle, S.D., 2002. Ecotone or ecocline: Ecological boundaries in  
1363  
1364 estuaries. *Est. Coastal Shelf Sci.* 55, 929-936.  
1365 548

1366  
1367 549 Banas, N.S., Hickey, B.M., 2005. Mapping exchange and residence time in a model of  
1368  
1369 Willapa Bay, Washington, a branching, macrotidal estuary. *J. Geophysical Res.-*  
1370 *Oceans* 110( C11), 10.1029/2005JC002950.  
1371 551

1372  
1373 552 Banas, N.S., Hickey, B.M., MacCready, P., Newton, J.A., 2004. Dynamics of Willapa  
1374  
1375 Bay, Washington: A highly unsteady, partially mixed estuary. *J. Phys. Oceanogr.* 34,  
1376 2413–2427.  
1377 554

1378  
1379 555 Banas, N.S., Hickey, B.M., Newton, J.A., Ruesink, J.L., 2007. Tidal exchange, bivalve  
1380  
1381 grazing, and patterns of primary production in Willapa Bay, Washington, USA. *Mar.*  
1382 *Ecol. Prog. Ser.* 341, 123–139.  
1383 557

1384  
1385 558 Banas, N.S., MacDonald, P.S., Armstrong, D.A., 2009. Green crab larval retention in  
1386  
1387 Willapa Bay, Washington: An intensive Lagrangian modeling approach. *Est. Coasts*  
1388 32, 893-905.  
1389 560

1390  
1391 561 Batchelor, G.K., 1967: *An Introduction to Fluid Dynamics*. Cambridge University Press,  
1392  
1393 615 pp.  
1394 562  
1395  
1396  
1397  
1398  
1399  
1400

- 1401  
1402  
1403 563 Bishop, M.J., Peterson, C.H., 2006. Direct effects of physical stress can be counteracted  
1404  
1405 564 by indirect benefits: oyster growth on a tidal elevation gradient. *Oecologia* 147, 426-  
1406  
1407 565 433.  
1408  
1409 566 Brown, J.R., Hartwick, E.B., 1988. Influences of temperature, salinity and available food  
1410  
1411 567 upon suspended culture of the Pacific oyster, *Crassostrea gigas*: I. absolute and  
1412  
1413 568 allometric growth. *Aquaculture* 70, 231–251.  
1414  
1415 569 Calbet, A., Landry, M.R., 2004. Phytoplankton growth, microzooplankton grazing, and  
1416  
1417 570 carbon cycling in marine systems. *Limnol. Oceanogr.* 49, 51-57.  
1418  
1419 571 Dame, R., Prins, T.C., 1998. Bivalve carrying capacity in coastal ecosystems. *Aquatic*  
1420  
1421 572 *Ecol.* 31, 409–421.  
1422  
1423 573 Dettmann, E.H., 2001. Effect of water residence time on annual export and  
1424  
1425 574 denitrification of nitrogen in estuaries: A model analysis. *Estuaries* 24, 481–490.  
1426  
1427 575 Dittman, S., 2000. Zonation of benthic communities in a tropical tidal flat of north-east  
1428  
1429 576 Australia. *J. Sea Res.* 43, 33-51.  
1430  
1431 577 Dumbauld, B.R., McCoy, L.M., 2015. Effect of oyster aquaculture on seagrass *Zostera*  
1432  
1433 578 *marina* at the estuarine landscape scale in Willapa Bay, Washington (USA).  
1434  
1435 579 *Aquaculture Environment Interactions* 7, 29-47.  
1436  
1437 580 Feldman, K.L., Armstrong, D.A., Dumbauld, B.R., DeWitt, T.H., Doty, D.C., 2000.  
1438  
1439 581 Oysters, crabs, and burrowing shrimp: Review of an environmental conflict over  
1440  
1441 582 aquatic resources and pesticide use in Washington State's (USA) coastal estuaries.  
1442  
1443 583 *Estuaries* 23, 141–176.  
1444  
1445  
1446  
1447  
1448  
1449  
1450  
1451  
1452  
1453  
1454  
1455  
1456

- 1457  
1458  
1459 584 Gangnery, A., Chabirand, J.M., Lagarde, F., Le Gall, P., Oheix, J., Bacher, C., Buestel,  
1460  
1461 585 D., 2003. Growth model of the Pacific oyster, *Crassostrea gigas*, cultured in Thau  
1462  
1463 586 Lagoon (Mediterranee, France). *Aquaculture* 215, 267–290.  
1464  
1465  
1466 587 Gao, S. 2019., Geomorphology and sedimentology of tidal flats. Pp 359-381 in (Perillo,  
1467  
1468 588 G.M.E., Wolanski, E., Cahoon, D.R., Hopkinson, C.S., eds) *Coastal Wetlands: an*  
1469  
1470 589 *integrated ecosystem approach*, 2<sup>nd</sup> ed. Elsevier.  
1471  
1472 590 Grangeré, K., Lefebvre, S., Bacher, C., Cugier, P., Ménesguen, A., 2010. Modelling the  
1473  
1474 591 spatial heterogeneity of ecological processes in an intertidal estuarine bay: dynamic  
1475  
1476 592 interactions between bivalves and phytoplankton. *Mar. Ecol. Prog. Ser.* 415, 141-  
1477  
1478 593 158.  
1479  
1480 594 Grizzle, R., Greene, J., Coen, L., 2008. Seston removal by natural and constructed  
1481  
1482 595 intertidal Eastern oyster (*Crassostrea virginica*) reefs: A comparison with previous  
1483  
1484 596 laboratory studies, and the value of in situ methods. *Est. Coasts* 31, 1208–1220.  
1485  
1486  
1487 597 Harrison, S.J., Phizacklea, A.P., 1987. Temperature fluctuation in muddy intertidal  
1488  
1489 598 sediments, Forth Estuary, Scotland. *Est. Coastal Shelf Sci.* 24, 279-288.  
1490  
1491 599 Hedgpeth, J.W., Obrebski, S., 1981. Willapa bay: a historical perspective and a rationale  
1492  
1493 600 for research. Office of Biological Services, US Fish and Wildlife Service,  
1494  
1495 601 Washington, D.C. FWS/OBS-81/03. 52 pp.  
1496  
1497 602 Herman, P.M.J., Middleburg, J.J., Widdows, J., Lucas, C.H., Heip, C.H.R., 2000. Stable  
1498  
1499 603 isotopes as trophic tracers: combining field sampling and manipulative labeling of  
1500  
1501 604 food resources for macrobenthos. *Mar. Ecol. Prog. Ser.* 204, 79-92.  
1502  
1503  
1504 605 Hickey, B.M., Banas, N.S., 2003. Oceanography of the U.S. Pacific Northwest coast and  
1505  
1506 606 estuaries with application to coastal ecology. *Estuaries* 26, 1010-1031.  
1507  
1508  
1509  
1510  
1511  
1512

- 1513  
1514  
1515 607 Hickey, B.M., Zhang, X., Banas, N., 2002. Coupling between the California Current  
1516  
1517 608 System and a coastal plain estuary in low riverflow conditions. J. Geophys. Res.  
1518  
1519 609 107(C10), 1029/1999JC000160.  
1520  
1521  
1522 610 Hsu, K., Stacey, M.T., Holleman, R.C., 2013. Exchange between an estuary and an  
1523  
1524 611 intertidal marsh and slough. Est. Coasts 36, 1137-1149.  
1525  
1526 612 Kang, C.K., Kim, J.B., Lee, K.S., Kim, J.B., Lee, P.Y., Hong, J.S., 2003. Trophic  
1527  
1528 613 importance of benthic microalgae to macrozoobenthos in coastal bay systems in  
1529  
1530 614 Korea: dual stable C and N isotope analyses. Mar. Ecol. Prog. Ser. 259, 79-92.  
1531  
1532 615 Kraus, N.C., 2000. Study of navigation channel feasibility, Willapa Bay, Washington.  
1533  
1534 616 Final Report. US Army Corps of Engineers Seattle District, Seattle. ERDC/CHL TR,  
1535  
1536 617 00-06.  
1537  
1538  
1539 618 Lawrence, D.R., Scott, G.I., 1982. The determination and use of condition index in  
1540  
1541 619 oysters. Estuaries 5, 23-27.  
1542  
1543 620 Lenihan. H.S., 1999. Physical-biological coupling on oyster reefs: How habitat structure  
1544  
1545 621 influences individual performance. Ecol. Monogr. 69, 251– 275.  
1546  
1547 622 Lenihan, H.S., Peterson, C.H., Allen, J.M., 1996. Does flow speed also have a direct  
1548  
1549 623 effect on growth of active suspension-feeders: An experimental test on oysters.  
1550  
1551 624 Limnol. Oceanogr. 41, 1359–1366.  
1552  
1553 625 Lomovasky, B.J., Brey, T., Kluegel, A., Iribarne, O., 2018. Distribution pattern, density  
1554  
1555 626 and growth of the stout razor clam *Tagelus plebeius* in a South-west Atlantic  
1556  
1557 627 estuarine system. J. Mar. Biol. Assoc. UK 98, 485-494.  
1558  
1559  
1560  
1561  
1562  
1563  
1564  
1565  
1566  
1567  
1568

- 1569  
1570  
1571 628 MacDonald, D.G., 2006. Estimating an estuarine mixing and exchange ratio from  
1572  
1573 629 boundary data with application to Mt. Hope Bay (Massachusetts/Rhode Island). Est.  
1574  
1575 630 Coastal Shelf Sci. 70, 326-332.  
1576  
1577  
1578 631 Muschenheim, D.K., Newell, C.R., 1992. Utilization of seston flux over a mussel bed.  
1579  
1580 632 Mar. Ecol. Prog. Ser. 85, 131-136.  
1581  
1582 633 Newell, C.R., 1990. Effects of mussel (*Mytilus edulis*, Linnaeus, 1758) position in  
1583  
1584 634 seeded bottom patches on growth at subtidal lease sites in Maine. J. Shellfish Res. 9,  
1585  
1586 635 113-118.  
1587  
1588 636 Oey, L.Y., 2005. A wetting and drying scheme for POM. Ocean Modelling 9, 133-150.  
1589  
1590 637 Peterson, C.H., 1991. Intertidal zonation of marine invertebrates in sand and mud.  
1591  
1592 638 American Scientist 79, 236-249  
1593  
1594 639 Peterson. C.H., Black, R., 1987. Resource depletion by active suspension feeders on  
1595  
1596 640 tidal flats -influence of local density and tidal elevation. Limnol. Oceanogr. 32, 143-  
1597  
1598 641 166.  
1599  
1600 642 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., R Core Team., 2016. nlme: Linear and  
1601  
1602 643 Nonlinear Mixed Effects Models. R package version 3.1-128, [http://CRAN.R-](http://CRAN.R-project.org/package=nlme)  
1603  
1604 644 [project.org/package=nlme](http://CRAN.R-project.org/package=nlme).  
1605  
1606 645 Pokavanich, T., Alosairi, Y., 2014. Summer flushing characteristics of Kuwait Bay. J.  
1607  
1608 646 Coastal Res. 30, 1066-1073.  
1609  
1610 647 R Core Team, 2015. R: A Language and Environment for Statistical Computing. R  
1611  
1612 648 Foundation for Statistical Computing, Vienna, Austria.  
1613  
1614 649 Ren, J.S., Ross, A.H., 2001. A dynamic energy budget model of the Pacific oyster  
1615  
1616 650 *Crassostrea gigas*. Ecol. Modelling 142, 105-120.  
1617  
1618  
1619  
1620  
1621  
1622  
1623  
1624

- 1625  
1626  
1627 651 Roegner, G.C., Hickey, B.M., Newton, J.A., Shanks, A.L., Armstrong, D.A., 2002.  
1628  
1629 652 Wind-induced plume and bloom intrusions into Willapa Bay, Washington. *Limnol.*  
1630  
1631 653 *Oceanogr.* 47, 1033–1042.  
1632  
1633 654 Ruesink, J.L., Roegner, G.C., Dumbauld, B.R., Newton, J.A., Armstrong, D.A., 2003.  
1634  
1635 655 Contributions of coastal and watershed energy sources to secondary production in a  
1636  
1637 656 northeastern pacific estuary. *Estuaries* 26, 1079–1093.  
1638  
1639 657 Ruesink, J.L., Yang, S., Trimble, A.C., 2015. Variability in carbon availability and  
1640  
1641 658 eelgrass (*Zostera marina*) biometrics along an estuarine gradient in Willapa Bay,  
1642  
1643 659 Washington, USA. *Est. Coasts* 38, 1908-1917.  
1644  
1645 660 Ruesink, J.L., Donoghue, C.R., Horwith, M.J., Lowe, A.T., Trimble, A.C., 2019.  
1646  
1647 661 Comparison of shallow-water seston among biogenic habitats on tidal flats. *PeerJ* 7,  
1648  
1649 662 e6616, <https://doi.org/10.7717/peerj.6616>.  
1650  
1651 663 Ruiz, C., Abad, M., Sedano, F., Garcia-Martin, L.O., Sanchez Lopez, J.L., 1992.  
1652  
1653 664 Influence of seasonal environmental changes on the gamete production and  
1654  
1655 665 biochemical composition of *Crassostrea gigas* (Thunberg) in suspended culture in El  
1656  
1657 666 Grove, Galicia, Spain. *J. Exp. Mar. Biol. Ecol.* 155, 249–262.  
1658  
1659 667 Ryu, J., Khim, J.S., Choi, J.W., Shin, H.C., An, S., Park, J., Kang, D., Lee, C.H., Koh,  
1660  
1661 668 C.H., 2011. Environmentally associated spatial changes of a macrozoobenthic  
1662  
1663 669 community in the Saemangeum tidal flat, Korea. *J. Sea Res.* 65, 390-400.  
1664  
1665 670 Senechal, J., Grant, J., Archambault, M.C., 2008. Experimental manipulation of  
1666  
1667 671 suspended culture socks: Growth and behavior of juvenile mussels (*Mytilus* spp.). *J.*  
1668  
1669 672 *Shellfish Res.* 27, 811–826.  
1670  
1671  
1672  
1673  
1674  
1675  
1676  
1677  
1678  
1679  
1680

- 1681  
1682  
1683 673 Simpson, J.H., Berx, B., Saurel, C., 2007. The interaction of tidal advection, diffusion  
1684  
1685 674 and mussel filtration in a tidal channel. *J. Mar. Systems* 68, 556–568.  
1686  
1687 675 Sullivan, J.C., Torres, R., Garrett, A., Blanton, J., Alexander, C., Robinson, M., Moore,  
1688  
1689 676 T., Amft, J., Hayes, D., 2015. Complexity in salt march circulation for a  
1690  
1691 677 semienclosed basin. *J. Geophys. Res.-Earth Surface* 120, 1973-1989.  
1692  
1693 678 Tomiyama, T., Komizunai, N., Ito, K., Omori, M., 2010. Spatial variation in the  
1694  
1695 679 abundance and condition of the bivalve *Nuttallia olivacea* in relation to  
1696  
1697 680 environmental factors and sublethal predation. *Mar. Ecol. Prog. Ser.* 406, 185-196.  
1698  
1699 681 Tweddle, J.F., Simpson, J.H., Janzen, C.D., 2005. Physical controls of food supply to  
1700  
1701 682 benthic filter feeders in the Menai Strait. *Mar. Ecol. Prog. Ser.* 289, 79-88.  
1702  
1703 683 Tweedley, J.R., Warwick, R.M., Potter, I.C., 2016. The contrasting ecology of temperate  
1704  
1705 684 macrotidal and microtidal estuaries. *Oceanogr. Mar. Biol. Ann. Rev.* 54, 73-171.  
1706  
1707 685 van Oevelen, D., Soetart, K., Middelburg, J.J., Herman, P.M.J., Moodley, L., Hamels, I.,  
1708  
1709 686 Moens, T., Heip, C.H.R., 2006. Carbon flows through a benthic food web:  
1710  
1711 687 Integrating biomass, isotope and tracer data. *J. Mar. Res.* 64, 453-482.  
1712  
1713 688 Viechtbauer, W., 2017. Package ‘metafor’. <http://www.metafor-project.org>  
1714  
1715 689 Wagner, R.J., Boulger, R.W. Jr., Oblinger, C.J., Smith, B.A., 2006. Guidelines and  
1716  
1717 690 standard procedures for continuous water-quality monitors—Station operation,  
1718  
1719 691 record computation, and data reporting: U.S. Geological Survey Techniques and  
1720  
1721 692 Methods 1–D3, 51 pp., <http://pubs.water.usgs.gov/tm1d3>.  
1722  
1723 693 Walles, B., Smaal, A.C., Herman, P.M.J., Ysebaert, T., 2016. Niche dimension differs  
1724  
1725 694 among life-history stages of Pacific oysters in intertidal environments. *Mar. Ecol.*  
1726  
1727 695 *Prog. Ser.* 562, 113-122.  
1728  
1729  
1730  
1731  
1732  
1733  
1734  
1735  
1736

- 1737  
1738  
1739 696 Warner, J.C., 2010. Using a composite grid approach in a complex coastal domain to  
1740  
1741 697 estimate estuarine residence time. Computers and Geosciences 36, 921–935.  
1742  
1743 698 Welschmeyer, N.A., 1994. Fluorometric analysis of Chlorophyll-a in the presence of  
1744  
1745 699 Chlorophyll-b and pheopigments. Limnol. Oceanogr. 39, 1985–1992.  
1746  
1747  
1748 700 Wheat, E.E., Ruesink, J.L., 2013. Commercially-cultured oysters (*Crassostrea gigas*)  
1749  
1750 701 exert top-down control on intertidal pelagic resources in Willapa Bay, Washington,  
1751  
1752 702 USA. J. Sea Res. 81, 33-39.  
1753  
1754 703 Wiberg, P.L., Carr, J.A., Safak, I., Anutaliya, A., 2015. Quantifying the distribution and  
1755  
1756 704 influence of non-uniform bed properties in shallow coastal bays. Limnol. Oceanogr.-  
1757  
1758 705 Methods 13, 746-762.  
1759  
1760 706 Whyte, J.N.C., Englar, J.R., Carswell, B.L., 1990. Biochemical composition and energy  
1761  
1762 707 reserves in *Crassostrea gigas* exposed to different levels of nutrition. Aquaculture  
1763  
1764 708 90, 157–172.  
1765  
1766  
1767 709 Xue, H., Du, Y., 2010. Implementation of a wetting-and-drying model in simulating the  
1768  
1769 710 Kennebec Androscoggin plume and the circulation in Casco Bay. Ocean Dynamics  
1770  
1771 711 60, 341–357.  
1772  
1773 712 Zu Ermgassen, P.S.E., Gray, M.W., Langdon, C.J., Spalding, M.D., Brumbaugh, R.D.,  
1774  
1775 713 2013. Quantifying the historic contribution of *Olympia* oysters to filtration in Pacific  
1776  
1777 714 Coast (USA) estuaries: implications for restoration objectives. Aquatic Ecol. 47,  
1778  
1779 715 149-161.  
1780  
1781  
1782  
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1795 717 FIGURE LEGENDS  
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1797 718 Figure 1. Two mechanisms for the depletion of phytoplankton from a parcel of water by  
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1799 719 intertidal suspension-feeders. In the familiar case (a), the parcel is depleted during its  
1800  
1801 720 passage across the intertidal zone on a single flood tide. In the case discussed in this  
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1803 721 paper (b), the parcel is depleted over several successive passages across the intertidal  
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1805 722 zone, returning due incomplete mixing in the channel at low slack tide. Parcels of water  
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1807 723 can be considered to have a residence time exceeding one tidal cycle. The length of time  
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1809 724 available for suspension-feeders to influence a parcel of water is controlled by the tidally  
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1811 725 averaged residual circulation (dotted arrow), rather than the tidal currents themselves.  
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1815  
1816 727 Figure 2. Water residence time on intertidal flats in Willapa Bay, Washington, from  
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1818 728 particle-tracking analysis of a 50 m-resolution numerical model with realistic intertidal  
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1820 729 bathymetry. Residence time (color scale, in units of tidal cycles) is the length of time  
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1822 730 that more than half of the 16 particles released in each 200 m square at high tide  
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1824 731 continue to be found in the intertidal zone at successive high tides, despite draining into  
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1826 732 the deeper channels in between. Solid contours give subtidal bathymetry at 5 m  
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1828 733 intervals. Box surrounds study site, as depicted in Fig. 3.  
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1830 734

1831  
1832 735 Figure 3. Tidal flat in Willapa Bay. The same area is depicted in (a) and (b), also  
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1834 736 corresponding to the boxed area in Fig. 2. (a) Aerial photograph overlaid with five  
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1836 737 stations where water properties and oyster performance were measured: South channel  
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1838 738 (ChS), North channel (ChN), Outer shore (ShOut), Middle shore (ShMid), and Inner  
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1840 739 shore (ShIn). Sensors were placed in 2007 for chlorophyll measurements (Chl) and in  
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1851 740 2008 for measurements of temperature and salinity (TS). Oysters were outplanted on  
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1853 741 tiles for 10 months. (b) Chlorophyll-a ( $\mu\text{g L}^{-1}$ ) along six transects one hour prior to  
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1855 742 afternoon high tide on 17 Aug 2008. The sensor recorded no data during three 50-sec  
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1857 743 periods, evident as gaps in transects.  
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1861  
1862 745 Figure 4. Temperature and salinity at the median water level on flood tides across a tidal  
1863  
1864 746 flat in Willapa Bay. (a) Salinity at near-channel stations linked to upwelling at  $48^{\circ}\text{N}$   
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1866 747  $125^{\circ}\text{W}$ . Upwelling is plotted with a 4-day lag, which represents the best correlation  
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1868 748 (Table 2). Salinity is a daily average of two flood tides. The portion of this summer-long  
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1870 749 time series from 4 to 15 August is shown in more detail, distinguishing each flood tide,  
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1872 750 in subsequent panels. (b) Salinity and (c) water temperature at a near-channel station and  
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1874 751 three stations toward shore. Station codes are in Fig. 3.  
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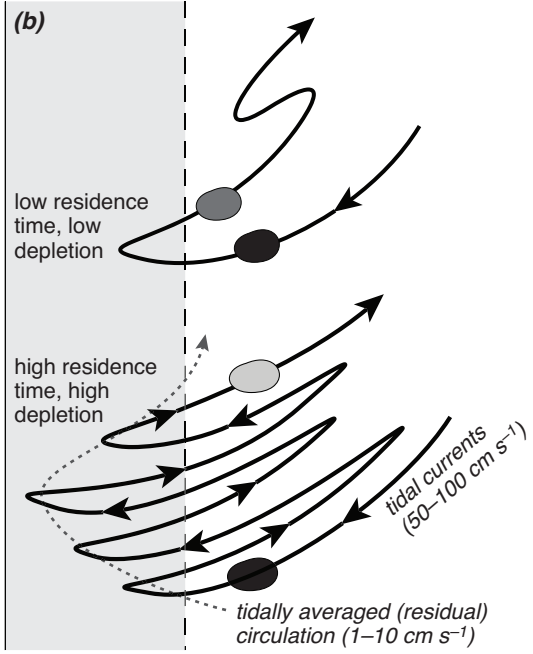
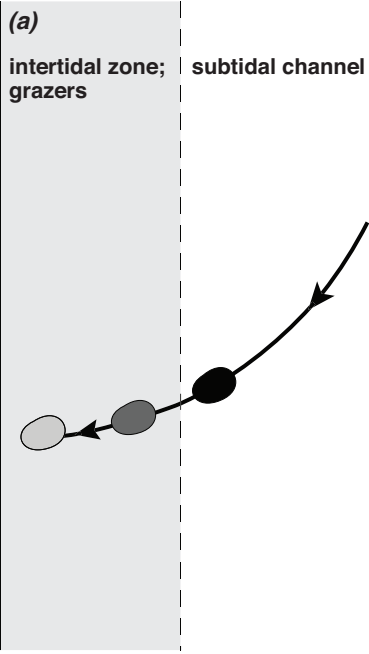
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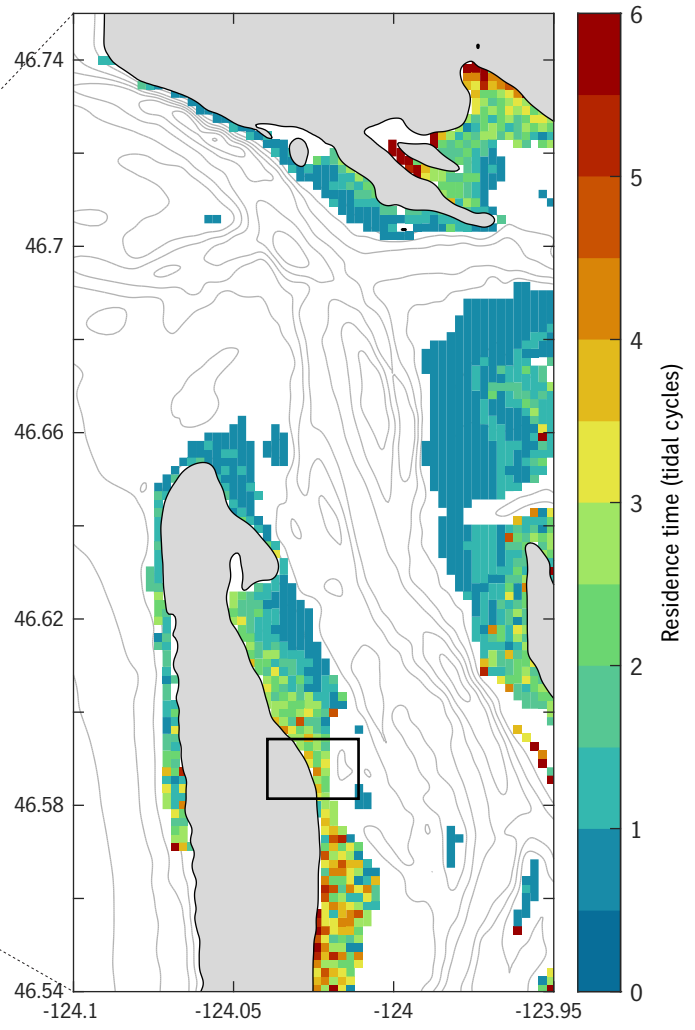
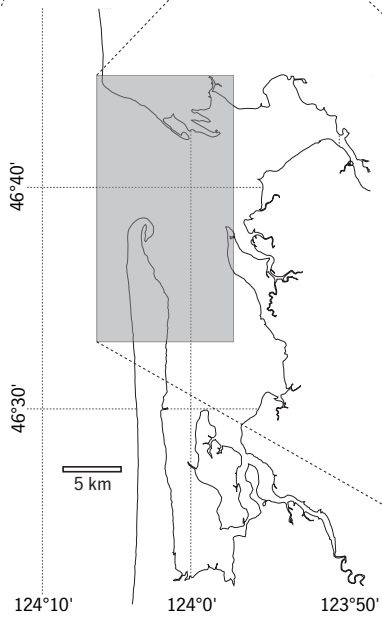
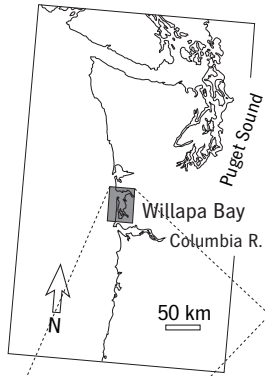
1878  
1879 753 Figure 5. Water column fluorescence at four stations across a tidal flat in Willapa Bay  
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1881 754 on each a) flood and b) ebb tide in 2007. Open symbols show values from one ShMid  
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1883 755 station where the sensor demonstrated strong drift during the last five days of  
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1885 756 deployment, and these values were not used in analyses. The following two panels show  
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1887 757 (c) fluorescence and (d) water temperature and salinity for six tidal cycles of data logged  
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1889 758 at 10-min intervals at one channel and one shore sensor during the early portion of the  
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1891 759 time series (2 Aug to 6 Aug 2007). Because fluorescence values exceeding  $25 \mu\text{g L}^{-1}$   
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1893 760 were censored, the data series is interrupted for ChN after the fourth tidal cycle, and no  
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1895 761 values were calculated around median water level. Station codes are in Fig. 3.  
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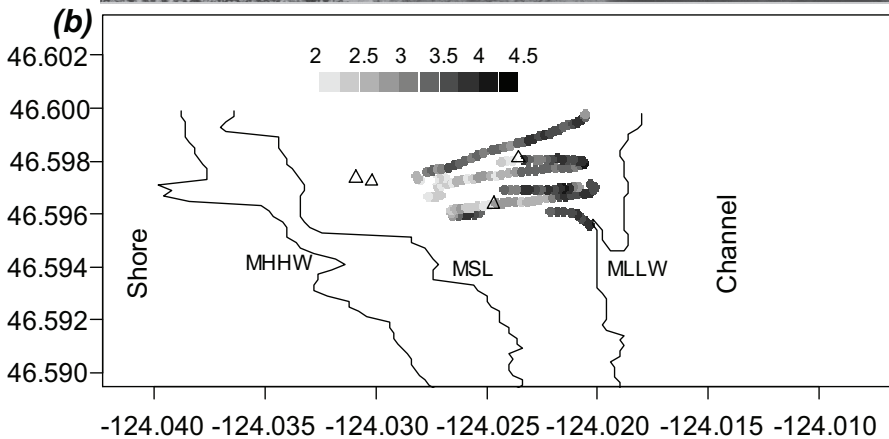
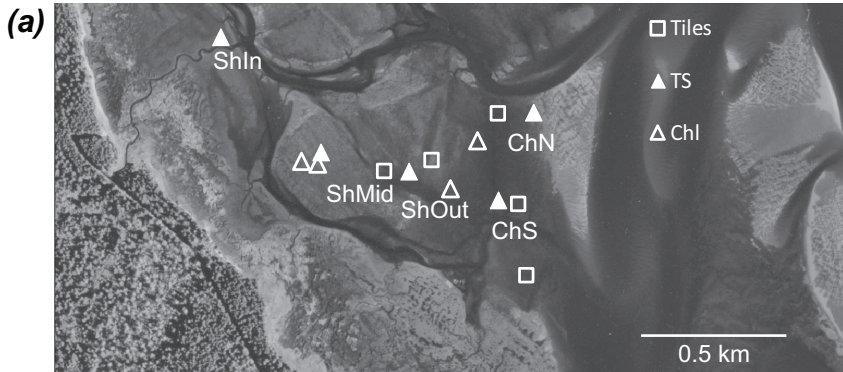
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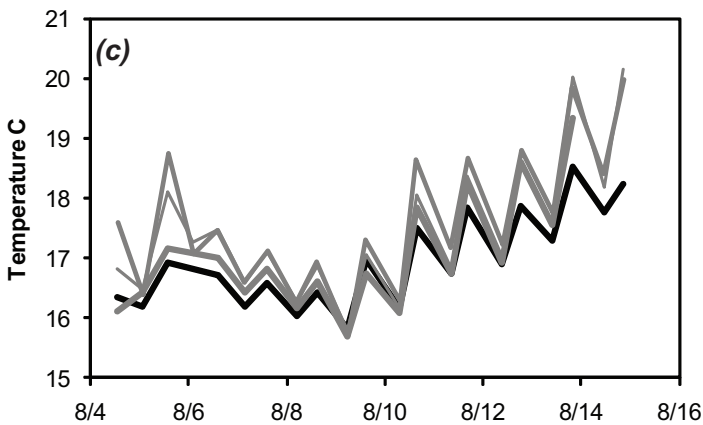
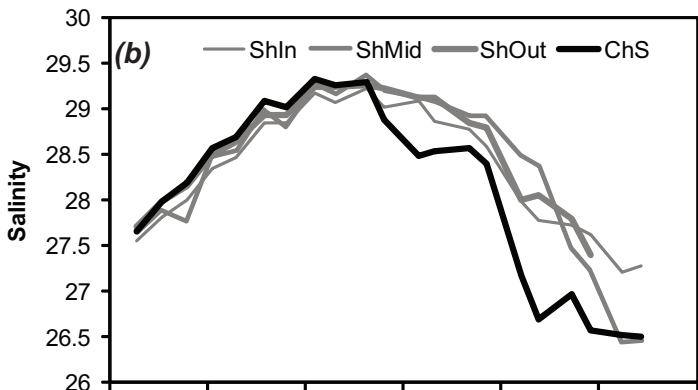
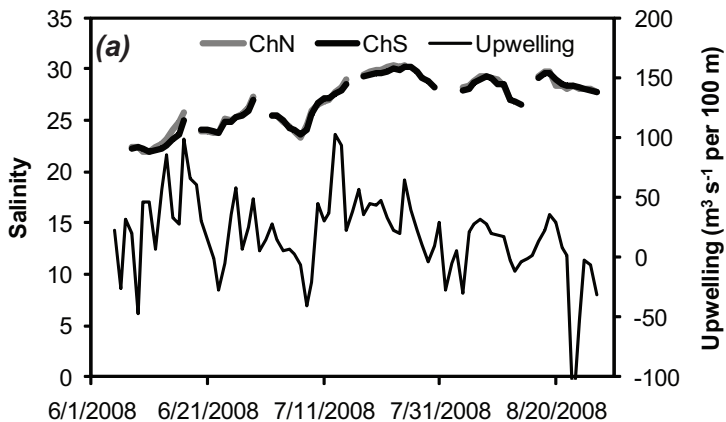
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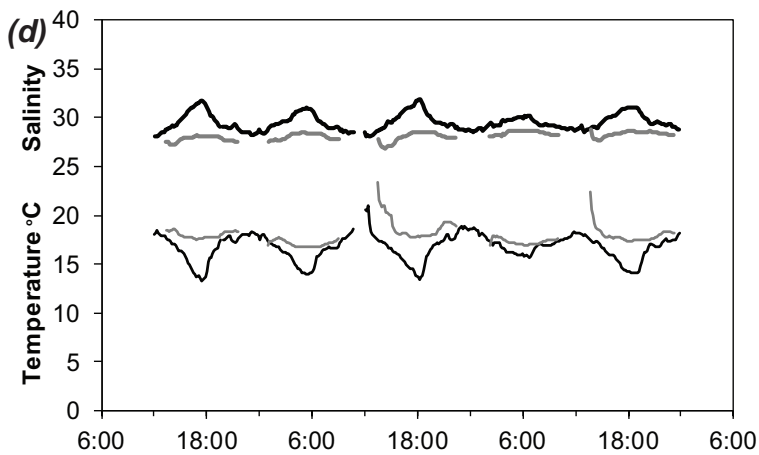
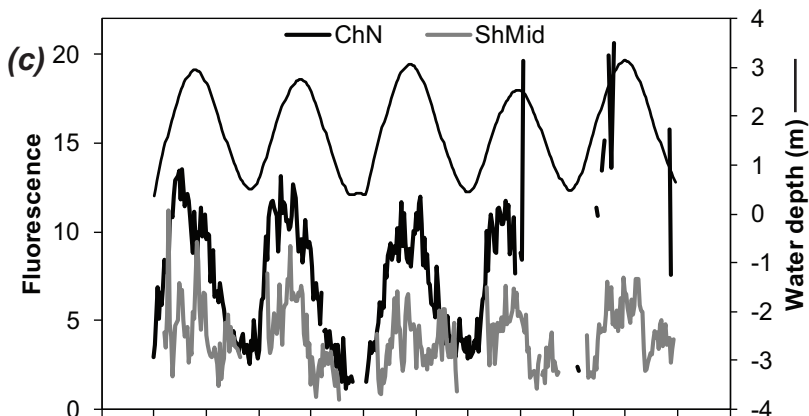
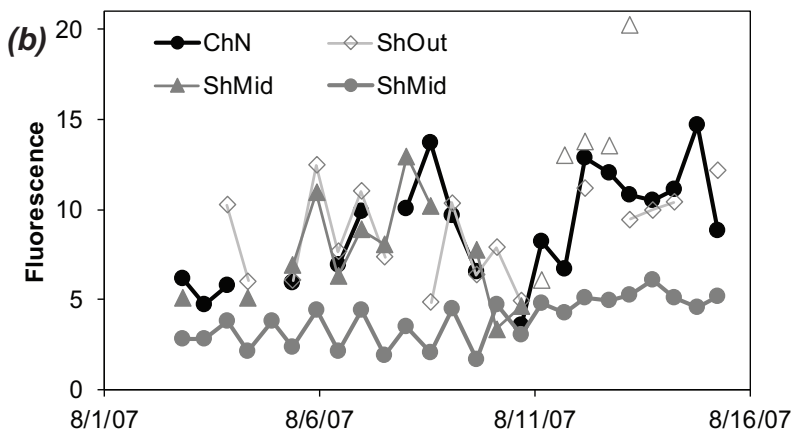
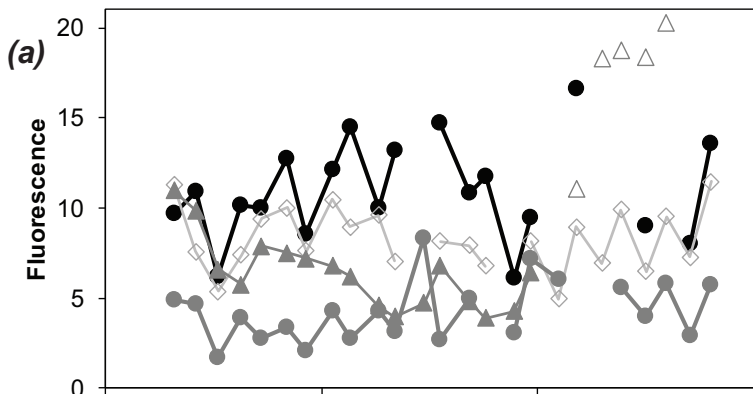
763 Figure 6. Performance of Pacific oysters (*Magallana = Crassostrea gigas*) across a tidal  
764 flat in Willapa Bay. A) Shell height, B) Condition, as dry flesh mass relative to shell  
765 height. Open points are individual oysters; dark points show means per tile. Oysters  
766 were outplanted at 1 cm in Aug 2007 and collected in May 2008. X-axis is reversed so  
767 that shore is to left and channel to right, to align with Fig. 2 and 3. Station codes are  
768 defined in Fig. 3.













Stations: ShMid ShOut ChN ChS

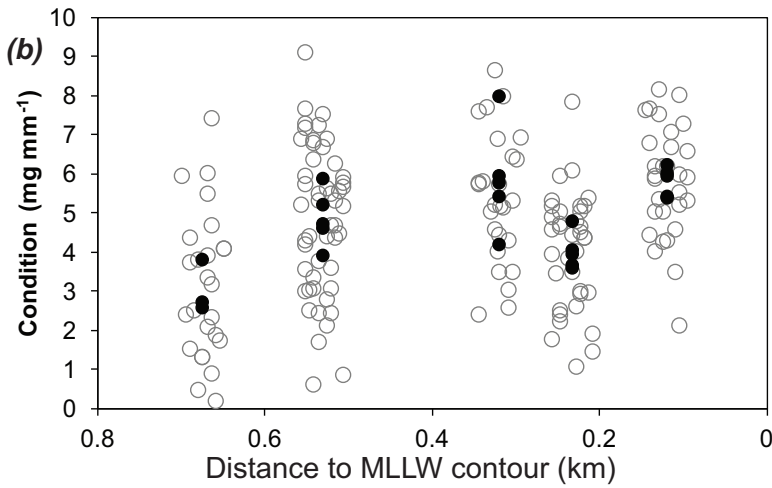
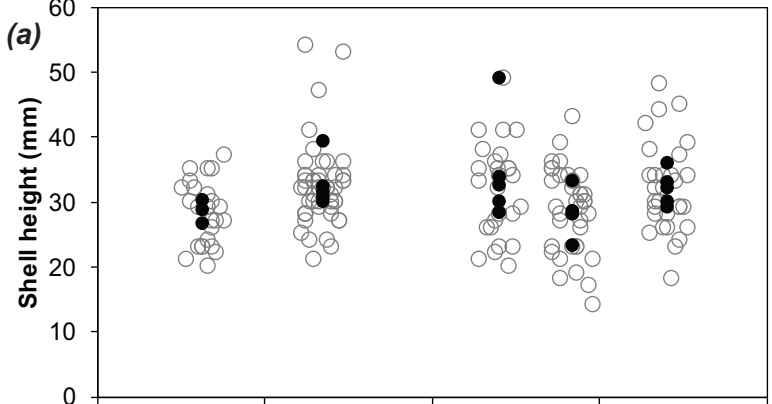


Table 1. Comparison of observed and modeled depth-averaged root mean square tidal velocity at six stations in the main channel and intertidal zone of Willapa Bay. Subtidal stations are at depths relative to mean sea level. Intertidal stations are at depths relative to mean high water.

Latitude	Longitude	Water depth	Depth-averaged rms velocity		
			Observed	Model	% error
<i>Subtidal (cf. Banas and Hickey, 2005)</i>					
46.694°N	124.097°W	10.1 m (MSL)	0.74 m/s	0.73 m/s	1%
46.697	124.064	10.1	0.78	0.81	4
46.696	123.973	9.75	0.61	0.78	28
46.521	123.999	12.2	0.53	0.63	19
<i>Intertidal (19-28 Jul 2008)</i>					
46.598	124.021	2.1 m (MHW)	0.21 m/s	0.22 m/s	5
46.598	124.031	2.0	0.074	0.10	35

Table 2. Model fit of linear models relating water properties measured from 8 Jun to 28 Aug 2008 at near-channel stations (ChN, ChS in Fig. 3) to coastal upwelling incorporating different lag times. Day of year was included as a predictor variable in all models to account for summer drought and seasonally-increasing salinity in Willapa Bay. Upwelling coefficients are estimates (standard error) from linear models, with significance as \*(<0.05), \*\*(<0.01), \*\*\*(<0.001). 63 daily values were available for ChN and 68 for ChS.

	Lag in days								
	No	No lag	1	2	3	4	5	6	7
upwelling									
ChN	0.617	0.614	0.641	0.669	0.711	<b>0.713</b>	0.692	0.685	0.667
adjusted r <sup>2</sup>									
ChN		0.004	0.013	0.018	0.022	0.023	0.021	0.020	0.016
Upwelling		(0.006)	(0.006)	(0.005)	(0.005)	(0.005)	(0.005)	(0.005)	(0.006)
coefficient			*	**	***	***	***	***	**
ChS	0.633	0.627	0.638	0.660	0.696	<b>0.728</b>	0.707	0.688	0.669
adjusted r <sup>2</sup>									
ChS		0.0001	0.008	0.014	0.019	0.022	0.021	0.019	0.015
Upwelling		(0.006)	(0.006)	(0.005)	(0.005)	(0.005)	(0.005)	(0.005)	(0.005)
coefficient				*	***	***	***	***	**

Table 3. Regression results relating salinity and temperature at shoreward sensors to channel station (ChS) with the most complete time series from 8 Jun to 28 Aug 2008. Model fit is provided as adjusted  $r^2$  for lags of different numbers of tidal cycles. Sample size (N) declines by 3-6 for each lag, due to gaps during sensor cleaning, but results were similar when sample sizes were made equivalent within each sensor comparison. Mean differences come from paired t-tests (relative to ChS) without lags.

Station	N	Lagged tidal cycles					Mean difference (95% CI)
		No lag	1	2	3	4	
<i>Salinity</i>							
ShIn	75	0.947	<b>0.960</b>	0.949	0.892	0.837	0.0009 (-0.085, 0.087)
ShMid	103	0.942	<b>0.947</b>	0.946	0.936	0.928	-0.241 (-0.367, -0.115)
ShOut	130	<b>0.972</b>	0.968	0.956	0.930	0.907	-0.153 (-0.226, -0.080)
<i>Temperature</i>							
ShIn	75	<b>0.768</b>	0.083	0.626	0.013	0.486	0.580 (0.450, 0.711)
ShMid	103	<b>0.865</b>	0.501	0.735	0.408	0.638	0.622 (0.490, 0.754)
ShOut	130	<b>0.925</b>	0.504	0.790	0.407	0.907	0.298 (0.227, 0.369)

Table 4. Results of paired t-tests relating fluorescence at shoreward stations to channel station (ChN) from 2 Aug to 15 Aug 2007. Values for each flood and ebb tide were based on an average of seven values logged at 10-min intervals around the median water level. Mean difference between each sensor and ChN is negative when fluorescence was lower towards shore. Two separate loggers were deployed 50 m apart at the Middle Shore station.

	Flood tides		Ebb tides	
	Mean difference (95% CI)	N	Mean difference (95% CI)	N
ShMid	-6.7 (-8.1, -5.4)	18	-5.0 (-6.3, -3.7)	20
ShMid	-4.2 (-5.9, -2.5)	16	0.0 (-1.7, 1.7)	8
ShOut	-2.7 (-3.9, -1.6)	19	-0.1 (-2.0, 1.8)	13

Supplemental material for Wheat EE, Banas NS, Ruesink JL. 2019. Multi-day water residence time as a mechanism for physical and biological gradients across intertidal flats. *Estuarine Coastal and Shelf Science*

Table S1. Position of oysters and sensors (YSI 6600) deployed intertidally on a wide tidal flat in Willapa Bay, Washington, USA. Station codes are North Channel (ChN), South Channel (ChS), Outer Shore (ShOut), Middle Shore (ShMid), and Inner Shore (ShIn). Latitude and longitude use WGS84 datum, and UTM uses 10T grid.

Data type	Station	Latitude °N (UTM Northing)	Longitude °W (UTM Easting)	Start date	End date	Elevation m MLLW	Distance to MLLW contour m
Oyster performance	ShMid	46.59702 (5150897)	124.02746 (421302)	3 Aug 2007	19 May 2008	0.6	673
Oyster performance	ShOut	46.59743 (5160940)	124.02550 (421453)	3 Aug 2007	19 May 2008	0.6	530
Oyster performance	ChS	46.59322 (5160469)	124.02153 (421750)	3 Aug 2007	19 May 2008	0.6	118
Oyster performance	ChN	46.59913 (5161126)	124.02274 (421666)	3 Aug 2007	19 May 2008	0.6	318
Oyster performance	ChS	46.59522 (5160695)	124.02618 (421397)	3 Aug 2007	19 May 2008	0.6	231
Salinity, Temperature	ShOut	46.59701 (5160895)	124.02643 (421381)	8 Jun 2008	28 Aug 2008	0.05	600
Salinity, Temperature	ShMid	46.59770 (5160975)	124.03008 (421102)	8 Jun 2008	28 Aug 2008	0.5	880
Salinity, Temperature	ChN	46.59917 (5161129)	124.02123 (421781)	8 Jun 2008	28 Aug 2008	0.6	200
Salinity, Temperature	ShIn	46.60188 (5161444)	124.03427 (420787)	8 Jun 2008	28 Aug 2008	0.9	1200
Salinity, Temperature	ChS	46.59592 (5160770)	124.02272 (421663)	8 Jun 2008	28 Aug 2008	-0.2*	320
Chlorophyll	ShMid	46.59735 (5160937)	124.03089 (421039)	2 Aug 2007	15 Aug 2007	0.7	940
Chlorophyll	ShMid	46.59724 (5160924)	124.03019 (421093)	2 Aug 2007	15 Aug 2007	0.55	890
Chlorophyll	ShOut	46.59637 (5160822)	124.02469 (421513)	2 Aug 2007	15 Aug 2007	0.1	470
Chlorophyll	ChN	46.59811 (5161014)	124.02356 (421602)	2 Aug 2007	15 Aug 2007	-0.1	380

\*deployed for the first two weeks at +0.2 m MLLW

Table S2. Estimates for slope and intercept (regression coefficients, with standard error) relating water properties to distance from channel (in km). Six transects were sampled at 5-sec intervals on 17 Aug 2008 to measure surface water with a YSI 6600. N=Number of samples per transect. Intercept is at the start of each transect. Chlorophyll-a (Chl) values were adjusted from sensor fluorescence (0.319) based on calibration by bottle samples. Latitude and longitude use WGS84 datum, and UTM uses 10T grid; both are presented for the start and end of each transect.

Transect	Latitude N (UTM Northing)	Longitude W (UTM Easting)	Chl intercept	Chl slope km <sup>-1</sup>	Salinity intercept	Salinity slope km <sup>-1</sup>
1, n=14	46.59559 (5160731)	124.02041 (421840)	4.23 (0.11)	-0.7135 (0.3979)	29.64 (0.021)	-1.05 (0.073)
	46.59596 (5160778)	124.02653 (421372)				
2, n=26	46.59705 (5160893)	124.02017 (421860)	4.12 (0.13)	-2.7345 (0.4185)	29.60 (0.017)	-1.35 (0.058)
	46.59606 (5160790)	124.02665 (421362)				
3, n=20	46.59705 (5160893)	124.02017 (421860)	4.45 (0.14)	-2.755 (0.4600)	29.85 (0.035)	-1.73 (0.118)
	46.59669 (5160860)	124.02768 (421284)				
4, n=27, 19	46.59804 (5161003)	124.02067 (421823)	4.01 (0.15)	-3.974 (0.493)	29.48 (0.020)	-1.60 (0.095)
	46.59705 (5160900)	124.02721 (4212321)				
5, n=23	46.59804 (5161003)	124.02086 (421809)	4.44 (0.15)	-3.469 (0.453)	29.61 (0.017)	-1.32 (0.051)
	46.59744 (5160944)	124.02822 (421244)				
6, n=31	46.59980 (5161199)	124.02056 (421834)	4.17 (0.12)	-4.327 (0.3915)	NA	NA
	46.59752 (5160953)	124.02813 (421251)				