EVALUATION OF CONTAMINANTS IN AGRICULTURAL SOILS IN AN IRRIGATION DISTRICT IN COLOMBIA

Martínez-Mera Eliana Andrea[ab]*, Torregroza-Espinosa Ana Carolina[a], Crissien-Borrero Tito José[c], Marrugo-Negrete José Luis[d], González-Márquez Luis Carlos[e]

[a.] Universidad de la Costa CUC, Grupo de Investigaciones en Desarrollo Agroindustrial Sostenible. Calle 58#55-66. Barranquilla-Colombia. atorregr4@cuc.edu.co *Corresponding author: emartine40@cuc.edu.co.

[b.] Estudiante de Doctorado en Ciencias Biológico Agropecuarias. Universidad Autónoma de Nayarit. México.

[c.] Universidad de la Costa CUC, Grupo de investigación en Gestión Educativa. Calle 58#55-66. Barranquilla-Colombia. rectoria@cuc.edu.co

[d.] Universidad de Córdoba, Grupo de investigación en Aguas, Química Aplicada y Ambiental. Carrera 6 #76-103. Montería, Córdoba. jmarugo@correo.unicordoba.edu.co

[e.] Universidad Autónoma de Occidente, Unidad Regional Guasave, Departamento de Ingeniería y Tecnología. Avenida Universidad s/n, CP 81048. Sinaloa, México. luis.gonzalez@udo.mx

Abstract

This study evaluated the concentration and distribution of heavy metals (HM) (Cr, Ni, Pb, Cd, Hg, and Zn) and pesticides (organochlorine and organophosphorus) and the relationship of these pollutants with the physicochemical properties of agricultural soils in an Irrigation District (ID) in Colombia. Soils samples were analyzed for pH, humidity, organic matter, P total, N total, electric conductivity (EC), cation exchange capacity, and texture (% sand, clay and silt). Canonical correlation was used to determined relationship between soil properties and HM. Soil pollution were evaluated with geoaccumulation index (Igeo), contamination factor (CF), degree of contamination (Cdeg) and pollution load index (PLI). The results indicated that, in general, the soils had adequate physicochemical conditions for the establishment and development of crops. The presence of pesticides in the soils was not reported. However, concentrations HM was detected (Zn>Cr>Ni>Pb>Hg>Cd). The soil characteristics (silt, clay, pH and EC) contributed to explain HM concentrations. The Igeo indicated that the soils are heavily contaminated with Hg (3<Igeo<4). The CF was very high for Hg (>6). The Cdeg presented moderate to considerable variations (>6Cdeg<24). The PLI indicated that the soils are contaminated (1.308). The presence of HM may be associated with the agricultural and quarries activities carried out near the ID. The impact caused by high concentrations of HM can lead environmental, economic and social impacts in the study zone.

Key words: heavy metals, pesticides, contamination index, soil properties.

Introduction

Soil is the most important basic natural resource for the support of agricultural production systems (De Alba *et al*., 2003) and the maintenance of productivity in these ecosystems depends on their physicochemical and biological characteristics (García *et al*., 2012; Martínez-Mera *et al*., 2017). However, soil is very sensitive to environmental variations (Chen *et al*., 2010). The anthropogenic activities of mining, changes in the use of soils and, use of agrochemicals in conventional agriculture have altered physicochemical properties, decreased edaphic populations and increased concentrations of some pollutants (Jaurixje *et al*., 2013). In this sense, agricultural production systems are a source of pollutants and, according with the physicochemical characteristics of the soil, facilitate their transfer through soil-plant, soil-groundwater and surface-soil water (Kabata- Pendias, 2011).

The pollutans in agricultural soils include HM and pesticides (Marković *et al*., 2010). HM are naturally present in small quantities or traces in the Earth's crust, soils and plants. Many of them are essential for the growth and development of plants, animals and humans (Galán-Huertos and Romero-Baena, 2008; Marković *et al*., 2010). These natural concentrations can be affected by the implementation of synthetic fertilizers and pesticides, manure and conventional solid waste compost (Wu *et al*., 2012; Alloway, 2013). High concentrations of HM can affect the environmental health of ecosystems, placing biota at risk (Galán-Huertos and Romero-Baena, 2008). On the other hand, pesticides are used to combat, repel and/or prevent unwanted organisms (plants or animals) during agricultural production (Gilden *et al.,* 2010). In Colombia, the intensive use of these products has increased the degradation of agricultural soils (Silva and Correa, 2009). Chemical agriculture models that use synthetic products such as fertilizers and pesticides are an important source of pollutants in edaphic systems (Rueda-Saá *et al*., 2011; Jiao *et al*., 2012).

Several studies have been carried out on the evaluation of physicochemical characteristics in soils and the presence of contaminants (Silviera *et al*., 2003; Alloway, 2013; Simón *et al*., 2013). Particularly in Colombia, progress has been made mainly in the departments of Atlántico and Córdoba (Yacomelo, 2014; Roqueme *et al*., 2014; Marrugo-Negrete *et al.*, 2017). The state of physicochemical properties in soils determines the quality and health of the soil because processes of adsorption, transport and degradation of contaminants depend on these characteristics (Bautista-Cruz *et al*., 2004). In fact, the physicochemical properties of soils, such as pH, OM, texture, mineralogy of clays, potential oxide reduction, carbonates, salinity and iron and manganese oxides and hydroxides, allow for the precipitation, dissolution and solubility of metals and pesticides (Galán- Huertos and Romero-Baena, 2008).

The use and proper management of soils is the most important method to conserve the soil environment. Otherwise, processes are generated that cause, not only loss of productive capacity, but also negative environmental, social and economic impacts (UPRA, 2013). Therefore, in the present study, the concentration of pollutants (HM and pesticides) was evaluated, as well as the relationship with physicochemical characteristics of agricultural soils in the South of the Atlántico Department. Studies are needed for a baseline for the definition of sustainable management strategies in this type of ecosystem because Colombia does not have a regulatory framework for concentrations of contaminants in soils (Rueda-Saá *et al*., 2011).

Methodology

Description of the study area

The municipality of Repelon is located to the west of the department of Atlántico (10º 29'40" N and 75º 08'27" W), with a surface area covers 35,172 ha (10.6% of the total area of the department of Atlántico). The Guájaro reservoir (10º 42 'N and 75º 6' W) stands out, a lentic body that supply the community for drinking water and for different productive activities. In its beginnings, the reservoir was capable of storing about 400,000,000 m^3 of water, in an area of 16,000 ha with five meters of average depth. Currently, El Guájaro has an extension of 11,647 ha, a perimeter of 114.28 km and an effective volume of 240,000,000 m^3 (Figure 1) (IDEAM, 2017). There are quarries in the middle and North area of El Guájaro, from which is extracted mainly construction material (sand, clay and limestone for cement, gravel and stone) (Carrillo and Cajuste, 1995). The most important economic activity in the region is the agriculture (CRA, 2014). An irrigation system (Figure 1) was constructed to transport water to the crops. This ID covering an area 4,000 ha, and it is supplied by the waters from the El Guájaro reservoir through a catchment channel. The irrigation system consists of a pumping station that transports water to two channels. From there, the upper distribution channel (15 km long) and lower channel (12 km long) facilitate irrigation through gravity (CRA, 2014). In this area, the main crops are, i.e. cotton, tomato, corn, sorghum, cassava, banana, rice, guava, papaya, and mango. The agricultural activity of transitory crops intensifies during the months of August-January (Alcaldía de Repelón, 2016).

As for the hydrometeorological conditions of the area, the temperatures are between 28 and 32°C. Rainfall varies according to the time of year: during the dry season (January-July), the average is 39 mm and, in the rainy season, it is 117.2 mm (August-December), with an annual average of 50 mm (Climate, 2017). However, the hydrometeorological conditions of the area result from the marked interannual variability, produced by the climatic anomalies El Niño and La Niña (Ruíz -Cabarcas and Pabón-Caicedo, 2013).

In terms of the general characteristics of the agricultural soils in southern Atlántico, the Repelón ID is located in quaternary deposits that occupy areas with a flat relief. The western side of the southern of Atlántico department has 50% of soils formed from clay sedimentary materials that are low in evolution, superficial, well drained, low to moderate in fertility and susceptible to erosion. These soils according to the texture, are classified as sandy loam clay and sandy loam at a depth of 0-30 cm and belong to the Inceptisol order. Table 1 describes the classification of the agricultural soils of the Repelón ID, suggesting a capacity for use, productivity and physicochemical characteristics (IGAC, 2008).

Table 1. Soil classification of the Repelón ID (IGAC, 2008).

Soil	Class I	Class II	Class III	Class IV
Topography	Plains, with deep parts	Moderately to strongly inclined, deep light erosion	Broken topography	Moderately steep, broken with moderate erosion
Drainage system	Moderate	Moderate-Good	Well drained	Moderate to excessive

Nitrogen	Poor	Poor	Poor	Poor
Phosphorus	High	High	Regular	Regular
Potassium	High	High	High	High
Salinity	Medium	Medium	Medium	Medium
pH	Neutral	Alkaline	Slightly alkaline	Slightly alkaline - acid
Crops	Tobacco, beans, corn	Grass	Grass	Grass
Observations	Improve drainages to avoid salinity in roots	Soils near the head. Heavy textures	Soils located in high parts of the municipality	No agricultural aptitude for high permeability is recommended crops of pasture or paddocks

Field phase

The field area was divided in three zones: northern zone (soil samples 1 to 4), central zone (soil samples 5 to 8) and, southern zone (soil samples 9 to 10). In areas with a history of agricultural activity, during the dry season a total of 10 topsoil samples (0–30 cm depth) were collected, which were obtained after the removal of the plant material from the soil surface, each soil sample comprised a composite of three subsamples. The characteristics and the locations of the sampling sites are shown in Table 2 and Figure 1. Samples for the physicochemical parameters and HM were stored in polyethylene bags. The samples for the pesticide analysis were wrapped in aluminum foil before being packed in polyethylene bags. All samples were transported under a controlled temperature (4 ± 1 °C) to the environmental laboratory of the Universidad de la Costa (CUC).

Table 2. Locations and description of the land uses during the sampling in the Repelón ID.

Soil Sampling	Longitud	Latitud	Types of Plantations
S1	-75.101639	10.509972	Forestry tree with grass
S2	-75.107222	10.519889	Fallow
S3	-75.112106	10.509928	Fallow
S4	-75.122989	10.517156	Annual crops
S5	-75.135569	10.488639	Annual crops
S6	-75.133444	10.483028	Annual crops

S7	-75.138222	10.469369	Fallow
S8	-75.135444	10.464306	Annual crops
S9	-75.142972	10.459889	Fruit tres, Fallow
S10	-75.132278	10.448222	Fallow

Laboratory phase

The physicochemical parameters were determined: moisture with the gravimetric method (IGAC, 2006), pH with the potentiometric method (NTC 5264, 2008), electrical conductivity ($EC_{1:5}$) with a conductivity meter (model EC300), total N (N total) with the Kjeldahl method (NTC 5889, 2011), organic matter (OM) from total organic carbon (Walkley-Black method) using a conversion factor (%C * 1.74) (IGAC, 2006), and cation exchange capacity (CEC) with the saturation method with ammonium acetate (NTC 5268, 2014). Additionally, the texture parameters (Bouyoucos-Densimeter Hydrometer) and total phosphorus (Pt) (Olsen *et al*., 1954) were determined in a certified laboratory (Zonas Costeras S.A.S).

Pesticide samples were analyzed with gas chromatography with an electron capture detector (CG-ECD), using EPA Method 8081B (US-EPA, 2007a) and 8141B (US-EPA, 2007b) for organochlorine and organophosphorus pesticides, respectively. The detection limit was 2 μg/Kg for organochlorine pesticides and 5 μg/Kg for organophosphorus pesticides. The quantification of HM zinc (Zn), nickel (Ni), cadmium (Cd), lead (Pb), and chromium (Cr), 0.5g of soil samples were digested with HNO_3/HCl 8:2 v/v in a microwave using EPA Method 3051A (US-EPA, 2007c). Additionally, concentrations of Hg heavy metal were evaluated using EPA Method 7471B, where 0.5g of soil sample was digested whit H_2SO_4/HNO_3 7.3v/v and 5% w/v $KMnO_4$ at 100 °C for 1h (US-EPA, 2007d). The detection limits (mg/Kg) varied for the metals (Zn = 5.0, Cr = 3.0, Ni = 20, Pb = 0.07, Cd = 0.002 and Hg = 0.001). The pesticides and heavy metals concentrations were calculated as the mean from triplicate determinations. These analyzes were carried out in the Toxicology and Environmental Management Laboratory of the Universidad de Córdoba-Colombia.

Quantification of soil pollution

Geoaccumulation Index (Igeo)

The Igeo for the metals were determined using:

$$\text{Igeo} = \text{Log}_2\left[\frac{\text{Cn}}{1.5\ \text{Bn}}\right]$$

There Cn is the concentration of metal examined in soil samples and Bn is the geochemical background concentration of the metal (n). Factor 1.5 is the background matrix correction factor due to lithologic. This index ranges from <0 uncontaminated, 0 <Igeo <1 low contamination, 1 <Igeo <2 moderately contaminated, 2 <Igeo <3 highly contaminated, 3 <Igeo <4 heavily contaminated, 4 <Igeo <5 very heavily contaminated and >6 extremely contaminated (Müller, 1981).

Contamination Factor (CF)

The level of contamination of soil by metal is expressed in terms of a *CF* calculated as:

$$CF = \frac{\text{Cmetal}}{\text{C background value}}$$

Where Cmetal is the concentration of metal examined in soil samples and C background value is the geochemical background concentration of the metal. This index ranges from CF <1 refers to low contamination; 1 < CF <3 means moderate contamination; 3 < CF <6 indicates considerable contamination and CF > 6 is very high contamination. The CF is a single element index (Hakanson, 1980).

Contamination Degree (Cdeg)

The Cdeg represents the sum of all pollution factors for all elements examined in a given site and it is determined using:

$$\text{Cdeg} = \sum_{i-1}^{n} \text{CF}$$

Where n are all the metals evaluated. There are four classes: Cdeg <n (low degree of contamination), n <Cdeg <2n (moderate), 2n <Cdeg <4n (considerable) and Cdeg> 4n (very high degree of contamination) (Hakanson, 1980).

Pollution Load Index (PLI)

The PLI is calculated by obtaining the n-root from the CF that were obtained for all the metals studied (n). The PLI was determined using (Tomlinson *et al.*, 1980):

$$PLI = \sqrt[n]{CF1 * CF2 * ...CFn}$$

According to the CF value, the PLI was categorized as <1 uncontaminated and >1 contaminated; with this value the relationship with soil quality was determined (Iqbal *et al.*, 2016).

Taking in account that in Repelón Municipality (Colombia) does not have local information about values of background or natural concentration elements, the values of background were taken from the limits proposed by the NOAA, which presents screening concentrations for contaminants (Buchman, 2008).

Information Analysis

Multivariate cluster analysis is presented in a dendrogram for metals and soils. The normality of the data was evaluated prior using Shapiro-Wilk test (Yap and Sim, 2011). As HM concentrations and soil properties did not show normal distribution, values were log-transformed. With the log-transformed data a multiple comparison was evaluated using a canonical correlation analysis, and find out the contribution of physicochemical properties of soils and heavy metals. Statistical analysis was performed with R.

Results

Physicochemical parameters

The physicochemical characteristics of the soil samples are shown in Table 3. The soils showed differences between the sites sampled. The pH was found to be slightly acidic (6.4) to slightly alkaline (7.2). The soil moisture was low, between 0.91-5.99%. The organic matter and total phosphorus presented values between 2.90-6.45% and 76.2-113.0 mg/Kg, respectively, which were high (IGAC, 2008; Martínez *et al.*, 2008; Novello and Quintero, 2009). The total N of the soils varied from very low (0.09%) to low (> 0.15%) (Hoskins, 2017). On the other hand, the $EC_{1:5}$ presented values <0.35 in all of the soil samples, indicating that the soils do not have salinity problems (Andrades and Martínez, 2014). The CEC was high (25-40 meq/100g) (Castellanos, 2016). Finally, three texture groups were found: silty clay loam (14.3% sand, 45.8% silty, 49.9% clay) and clay (14% sand, 39.8% silty, 46.2% clay) in the northern zone; silty clay (9.6% sand, 44.3% silty, 46.1% clay) in the central zone; and clay (6.5% sand, 38.4% silty, 55.1% clay) in the southern zone.

Table 3. Physicochemical properties of the agricultural soils in the Repelón ID.

Soil sampling	**pH**	**Humidity (%)**	**Organic matter (%)**	**P (mg/Kg)**	**N (%)**	**EC (dS/m)**	**CEC (meq/100g)**	**Texture**
S1	7.2	4.25	2.90	76.2	0.15	0.10	35.4	sicl
S2	6.8	3.95	5.20	90.6	0.07	0.03	35.3	sicl
S3	6.9	5.77	4.05	94.4	0.11	0.01	41.4	c
S4	7.2	3.71	6.00	98.5	0.26	0.02	29.2	c
S5	7.2	1.48	5.60	101.6	0.12	0.07	39.3	sic
S6	7.2	4.81	5.97	113.0	0.11	0.02	48.8	sic
S7	6.5	0.91	3.42	106.3	0.26	0.04	39.1	sic
S8	7.2	5.99	6.45	108.5	0.30	0.01	47.2	sic
S9	6.4	2.43	3.83	102.8	0.21	0.01	54.4	c
S10	6.4	5.61	3.58	111.3	0.16	0.04	51.9	c

EC (Electric Conductivity); CEC (Cation Exchange Capacity); Texture: sicl (silty clay loam), c (clay), sic (silty clay).

Pesticides

The presence of organochlorine pesticides such as heptachlor benzene, aldrin, endosulfan, dieldrin, endrin, and 4,4'DDT and organophosphorus pesticides such as malathion, chlorpyrifos-methyl, fention, tridemofon, diazinon, cis-chlorfenvinfos and dimethoate was not reported. The results obtained were below the detection limit of the equipment (<2 μg/Kg for organochlorine pesticides and < 5 μg/Kg for organophosphorus pesticides).

Heavy metals

The concentrations of HM showed variations in the soil samples. The average levels of HM in the agricultural soils follow a decreasing order as: Zn> Cr> Ni> Pb> Hg>. The HM Zn, Cr, Ni and Cd showed a similar pattern of concentrations in the field area. In general, highest concentrations (Zn, Cr, Ni, Hg and Cd) were found in the southern area except for the Pb that showed high concentration in the central zone. Additionally, northern and central zones had little differences in concentrations for Zn, Cr, Ni and Cd. Similar pattern was showed for Pb in northern and southern zone. Finally, Hg showed variations between central (high concentration) and northern zone (low concentration) in comparison with southern zone.

On the other hand, the HM studied were highest than the world soil reference (NOAA) (Table 4). Concentrations of Hg were 8.5 times higher than reference, followed by Ni (3.3 times), Zn (2.01 times), Cr (1.84 times), Pb (0.34 times) and Cd (0.2 times) that was the HM in lowest concentration.

Table 4. Heavy metals concentrations in agricultural soils from Repelón (Colombia).

Soil sampling	**Heavy metals (mg/Kg)**					
	Zn	**Cr**	**Ni**	**Pb**	**Hg**	**Cd**
S1	94.1	68.5	44.2	4.94	0.09	0.33
S2	99.2	70.4	47.0	5.18	0.09	0.33
S3	91.5	73.8	45.2	5.19	0.10	0.29
S4	75.1	51.0	34.5	3.88	0.74	0.19
S5	94.5	68.0	40.0	4.70	0.90	0.30
S6	92.5	65.9	42.5	6.51	0.92	0.31
S7	109.3	70.6	42.6	5.69	0.10	0.24
S8	98.1	66.6	43.0	5.15	0.51	0.32
S9	103.3	71.6	46.4	6.14	0.73	0.43
S10	109.4	73.9	47.2	6.38	0.69	0.43
Mean	**96.7**	**68.0**	**43.3**	**5.38**	**0.49**	**0.32**
NOAA[1]	**48**	**37**	**13**	**16**	**0.05**	**1.6**

[1] Background (mg/Kg)

Geoaccumulation index (Igeo): The concentrations of Pb and Cd did not represent contamination in the agricultural soils of the Repelón ID. The HM Zn and Cr presented a slight degree of contamination, Ni had moderate contamination and Hg had a degree of strong contamination (Figure 2).

Contamination Factor (CF): The HM in the Repelón agricultural soils presented variations in the CF. Figure 3 shows the CF for the evaluated soils. The CF was low for Cd and Pb (0.20

and 0.33, respectively); moderate for Cr and Zn (1.84 and 2.04, respectively); considerable for Ni (3.33) and very high for Hg (8.45).

Contamination Degree (Cdeg): The Cdeg in the northern zone varied between moderate (9.56) to considerable (18.7) where the soil sample S4 presented high contamination degree in this zone. In the central zone, the Cdeg varied between considerable (21.0) to moderate (9.72), in this zone the soil sample S6 showed highest level of Cdeg of all soils. Finally, in southern zone the Cdeg was considerable (20.7).

Pollution Load Index (PLI): In this study, the evaluation of the overall toxicity status of all soil samples were PLI> 1, this value suggest that quality of the agricultural soil is deteriorating (Mir-Mohammad *et al.,* 2016). The pattern of pollution in each soil sample was similar to Cdeg, where soil samples S4 and S6 showed higher values, as the same manner in southern zone.

Multivariate cluster analysis

Dendrogram in Figure 4a. enabled the identification of two major clusters. Cluster 1 is composed of Pb, Hg and Cd. The second cluster is composed of Cr, Ni y Zn. Similarly, sampling points were also analyzed by clustering methods and organized in the dendrogram to identify similar groups (Figure 4b). The sampling sites could be grouped in three clusters, with the majority in cluster 1 with 7 samples, and cluster 2 with 2 samples. In addition, among all the sampling sites, 70% represent cluster 1, with a similar percentage of samples located in the north and center zone of the Repelón ID (Figure 4).

Canonical correlation analysis

The correlations between physical-chemical soil characteristics (G1) and HM concentrations (G2) in the soil samples are shown in Table 5. Six canonical correlations were obtained (L1-L6), of these correlations only L1, L2 and L3 were statistically significant (P-value <0.01). The high correlation coefficient (near 1) means a good relation between variables, if R^2 is more than 0.7, it can be strongly correlated. In L1, the variability (99%) was explained by silt, clay and Cd, they had the biggest positive coefficients. Additionally, Ni and Hg had a high negative standardized coefficient. In L2, the 99% of the variability was explained by silt, Hg, Zn and Cr. Whereas EC had a high negative standardized coefficient. Finally, in L3 the 95% of the variability was explained by pH, EC, Ni and Hg. Conversely, Cd, Cr and Pb had a high negative standardized coefficient.

Table 5. Canonical correlations between physical-chemical soil characteristics and heavy metals concentrations in the agricultural soils of the Repelón ID.

Groups	**Characteristics**		**Canonical correlations**					
			L1	**L2**	**L3**	**L4**	**L5**	**L6**
G1	pH		0.24	0.15	1.18	0.35	-0.05	0.18
	Humidity		-0.40	0.48	-0.15	-0.22	0.12	-0.66
	Organic matter		0.55	-0.20	-0.30	-0.34	-0.17	-0.52
	CEC		0.41	-0.92	0.88	1.42	-0.22	0.98
	EC		-0.38	-0.23	0.20	-0.83	-0.15	0.83
	Total N		0.16	0.00	0.37	-0.66	0.39	0.21
	Total P		0.19	-0.05	0.10	0.08	0.45	0.82
	Texture	Sand	0.36	0.49	0.04	2.27	0.80	1.29
		Silt	1.40	0.74	0.47	3.92	0.02	3.82
		Clay	1.12	0.12	0.36	1.27	-0.68	2.64
G2	Zn		-0.46	0.68	0.09	0.40	-1.67	-0.81
	Cr		0.09	0.65	-1.69	-1.05	1.90	-0.53
	Ni		-2.94	0.06	4.80	-1.38	0.64	-0.44
	Pb		0.59	-0.31	-1.19	0.14	-0.28	1.86
	Hg		-1.81	0.97	1.60	-0.28	1.08	-0.53
	Cd		3.16	-0.24	-1.86	1.40	-0.78	0.20
	R^2		0.99	0.99	0.95	0.82	0.39	0.04
	P-value		0.00	0.00	0.01	0.32	0.93	0.99

P-value<0.01 Significant correlation.

Discussion

The soils of the Repelón ID presented adequate physicochemical characteristics. The pH presented values in the range of 6.5-7.5. In general, the pH of the soils was found within the range of acceptable values for the development of crops and the availability of nutrients (Andrades and Martínez, 2014). The low moisture content of the soils was related to the sampling period (dry season). Similarly, the high OM% is associated with the fallow where the vegetation cover of the soils (grass), which facilitates the accumulation of organic waste and the low precipitation prevents water erosion (Guangwei *et al.*, 2006). The Ptotal content in the soil can be altered by the removal of crops (approximately 80% is absorbed by the plants), water erosion and OM mineralization (Suñer *et al*., 2001; Novello and Quintero, 2009). However, in the Repelón soils, these factors did not influence this nutrient. Conversely, the high content of Ptotal depend to the superficial horizon, where high concentrations of this element are found because plant residues accumulate on the surface (Novello and Quintero, 2009). The low Ntotal content was related to the variability in soil temperature and moisture. High temperatures and low rainfall influence microbial degradation. Therefore, these variables affect the N supply capacity of the soil. A low content of salts can depend, to a large extent, on agricultural practices; during the dry season, the application of fertilizers is very low. Finally, the high CEC was due to the absence of salts, the texture (high clay content) and the OM present in the soil (Hoskins, 2017).

In this study, the soils were class I and II due their topography (IGAC, 2008), making them suitable for agricultural activities, and the parameters Ntotal and Ptotal did not present variation in the values previously reported (low and high, respectively). However, parameters, such as pH and salinity, varied. The factors associated with changes in pH in the soil were possibly related to the agricultural history of the soils although, currently the economic agricultural activity is subject to the rainy season. The application of chemical inputs was related to changes in the pH in the soil (Martínez-Mera *et al*., 2017). Additonally, La Niña phenomenon where the soils were flooded, the soil pH increased with flooding time (Kashem and Singh, 2001). On the other hand, salinity in soils can be caused by natural or anthropic processes. In the former, the soluble salts are found in the subsoil and rock deposits. On the other way, the salinity generated by agricultural activities can be caused by inadequate irrigation methods and the handling of chemical substances (McKenzie, 2013). Taking these factors into account, the decrease in salinity in Repelón soils may be related to the decrease in agricultural activities associated with climate change (dry season) or improvement of agricultural practices over time.

Although the use of LorsbanTM 4E insecticide, glyphosate herbicide and NPK 15-15-15 (Fertilizer-Triple 15) (Martínez-Mera *et al*., 2017) has been reported in the Repelón ID when the agricultural activities increase during rainy season, the non-detection of organochlorine and organophosphorus pesticides could be related four reasons: *i*). low agricultural activity during the dry season, *ii*). climatic conditions of the municipality, *iii*). physicochemical properties of the evaluated soils, and *iv*). environmental regulation. In terms of climatic conditions characterized by high rainfall that in some periods have caused flooding (winter wave 2010) (Ruíz -Cabarcas and Pabón-Caicedo, 2013); long periods of drought conditions and high solar radiation, during 2016 the variation of rainy was 0.2 mm (January) and 256.11 mm (May) with temperature between 36 °C (March) and 32 °C (November) and average of ultraviolet index 7-9 (very high harmful ultraviolet radiation) (Word Weather, 2018); these conditions favor photolysis of pesticides and increased evaporation of volatile or semi-volatile substances (Narváez *et al*., 2012). Whereas, during periods of high rainfall, the soil becomes saturated, favoring the leaching of pesticides (Uzcátegui *et al*., 2011; Ruíz-Cabarcas and Pabón-Caicedo, 2013). In relation to physicochemical properties, such as high OM and clay, they have the ability to adsorb or immobilizing pesticides, leaving them unavailable for biodegradation (Cornejo and Jamet, 2000). On the other hand, the Colombian Agricultural Institute (ICA) in Resolution 2189 in 1974, prohibited the sale of fungicides that containing phenylmecury, chemical inputs were use over 50 years ago. Additionally, in Resolution 366 in 1987, and Resolutions 531, 540, 723, 724 and 874 in 1988, prohibited the sale of insecticides containing active ingredients such as aldrin, heptachlor, dieldrin, chlordane and camphechlor; it is likely that these pesticides characterized by their persistence, have degraded during the 28 years that have elapse and these regulations are being complied with. The commonly used pesticides are classified as organophosphates, with high toxicity and low chemical stability; therefore, they have high degradability (Silveira *et al.*, 2003). Taking into account the fact that the average life span of LorsbanTM 4E (30-60 days) and glyphosate (1-130 days) vary depending on the soil and climatic conditions, this condition is associated with the absence of pesticides in the analyzed soils because the highest frequency of application occurs during the rainy season.

The indicators to evaluate the contamination by HM presented variations, and showed that in general the ID is contaminated. Pollution sources are associated with natural phenomenon and anthropogenic activities. During 2010 the phenomenon of flooding soil for a long period could provide availability of HM due little aeration of soil, reduction conditions are favored, this condition increases the toxicity of some metals such as Mn (manganese), Cu (copper), Zn, Cd, and Cr (Poot *et al.*, 2007; Reichman, 2002). On the other hand, the anthropogenic activities where the municipality of Repelón is part of the Calamarí Mining District and has

27 active mining operations (where 9 are illegal), taking up approximately 50 ha in areas surrounding the El Guájaro reservoir that generate problems in different bodies of water (Alcaldía de Repelón, 2017). In the study area, there are quarries in the middle and north area of El Guájaro, from which is extracted mainly construction material (gravel, sand, stone and limestone), activity that possibly by erosive processes accumulate metals in particular areas (CRA-CARDIQUE, 2002; CRA, 2014), which could reach the El Guájaro reservoir and, later, the agricultural soils of the Repelón ID through the irrigation water. Belmonte-Serrato *et al.* (2010) and Vallejo *et al.* (2016), reported that the extraction of mineral resources (represented by quarries) contributes significantly to increase the concentrations of Zn, due it contributes 10-15% of the total sediments delivered by laminar erosion, which are possibly aggregated by atmospheric deposition and runoff (CRA, 2014). This result corroborate the high concentration of Zn in comparison with the others HM. Additionally, Torregroza-Espinosa *et al.* (2018), reported the presence of HM (Zn, Pb, and Hg) in water and surface sediments of El Guájaro Reservoir, the highest levels were found in the southwest and northern zones of this body of water, it is possible that the pollutants are associated with agricultural and mining activity, floodgate (El Guájaro Reservoir with Repelón ID) and the connection with the Canal del Dique (from Magdalena river) (Figure 1). On the other hand, industrial activities (aquaculture and shrimp farm), municipal waste disposal and agricultural activities, involving irrigation and applications of chemical substances, pesticides and manure, can also be sources of HM (Belmonte-Serrato *et al*., 2010; Marrugo-Negrete *et al.* 2017). The problems associated with irrigation systems depend on the quality of the water, which can be contaminated by the main tributary or receive wastewater discharge, providing heavy metals such as Ni, Pb, Cr, Cd, Zn, Cu, Hg, Mn, and Fe (iron), among others (Kim *et al*., 2015). The application of fertilizers that provide macronutrients (N, P, K) contain impurities of Cd and Pb, which can significantly increase their content in the soil because of regular use since they are fundamental for the growth of the plants. Likewise, phosphate fertilizers provide Cd, Hg, Pb, Co, Cu, As and Zn as impurities (Gimeneo-García *et al*., 1996; Wuana and Okieimen, 2011; Alloway, 2013). Other fertilizers, such as copper sulfate and iron sulphate, contain Pb and Ni (Gimeneo-García *et al*., 1996). Even fertilizers from animals, such as manure, can also provide As, Zn and Cu as a product of the animals' diet

(Andrades and Martínez, 2014). And some pesticides can also contribute As, Pb, Hg, Cu and Zn (Alloway, 2013; Gaw *et al.*, 2006).

During the sampling, intensive applications of chemical products in the evaluated soils were not reported. However, historically, the municipality of Repelón is known as the agricultural pantry of the Atlántico Department. Therefore, the presence of HM in the soils cannot just be associated with mining activities; it is likely that, during the winter season when agriculture intensifies, common practices such as the application of chemical inputs are a source of HM. Mining is responsible for soil degradation. Artisanal mining processes significantly alter the landscape and cause air, soil and water pollution. Quarries are associated with forced migration and loss of soil biodiversity, and destruction of fragile ecosystems (Vallejo *et al.*, 2016).

Dendrogram for heavy metals enabled the identification of two major clusters. Cluster 1 is composed with the heavy metals (Pb, Hg, Cd) that are related with anthropogenic activities like wastewater, waste combustion and agricultural activities (fertilization and pesticide application), specifically with the application of phosphate fertilizers (Marrugo *et al.,* 2017; Wuana and Okieimen, 2011). And in the Cluster 2 heavy metals (Zn, Cr, Ni) have relationship with the phenomenon of flooding soil generated by the floods of the Magdalena River or mining activities of quarries, as previously explained. In the dendrogram for sampling sites, two major clusters were observed. Cluster 2 shows the sampling site 1 to the sampling site 7, with the exception of sampling site 4, possibly grouped by agricultural activities. While the second cluster includes sites 9 and 10. These sampling sites are located in the southern zone of the ID of Repelón, and presented high values in the concentrations of all the metals evaluated. In this area, the Canal del Dique connects with El Guájaro Reservoir through two floodgates (Figure 1), where the exchange of 50,000 tons/year of sediment to the reservoir (CRA, 2014) which may explain why those soils are grouped. Similarly, that zone was the most affected by the floods caused by La Niña phenomenon.

In the present investigation, a positive correlation was observed between silt and clay with Cd, and a negative correlation between Ni and Hg, with silt and clay. In the second correlation, a positive relationship was observed between silt and Hg, Zn y Cr, and a negative relationship between CEC with Hg, Zn and Cr. In the third correlation a high negative correlation was found between the pH and the Cd, Cr and Pb concentrations, and a positive correlation between pH and Ni and Hg concentration. The soils properties play an important role reducing or increasing the availability and toxicity of metals (Galán-Huertos and Romero-Baena, 2008). In this sense, the positive correlation with silt and clay, may be explained by the presence of clay minerals and Fe and Mn oxides associated forming silt sized aggregates (Melo *et al.*, 2000). Additionally, clay presents adsorptive properties that enhanced the retention of metals that decreases its bioavailability (Jung, 2008; Marrugo *et al.*, 2017). The CEC is moderately related to the content of clay and OM, thus soils retain heavy metal cations (Ahumada *et al.*, 1999; Marrugo *et al.*, 2017). In agreement with IGAC (2008), the Caribbean soils evaluated presented high content of OM, the higher CEC has the greater the capacity of the soil to fix metals (Ahumada *et al.*, 1999). On the other hand, the negative correlation between CEC and metals concentrations, is possibly explained by the formation of chelate complexes with the OM, due the high CEC (in a function of the content of OM) influences the solubility and assimilation (Angelova *et al.*, 2013). The adsorption and exchange of metals is generally attributed to the properties of the adsorbent, the solvent, the concentration, the valence and the degree of hydration of the cations (Sparks, 2003). Ions with a smaller hydration radius can get closer to the adsorption surface, and adsorption of these can be favored. The average concentration of metals in the analyzed soils was Zn> Cr> Ni> Pb> Hg> Cd; such sequence is not related to the hydration radius of the metals Pb (0.401 nm) <Ni (0.404 nm) <Cd (0.426 nm) <Zn (0.43 nm) <Cr (0.461). Finally, the pH affects several mechanisms of metal retention in soils (Carrillo and Cajuste, 1995), and can be considered the most important parameter that influences the processes of sorption-desorption, precipitation and dissolution. Likewise, the formation of complexes and the reactions of oxide-reduction where the bioavailability of the elements is inversely proportional to soil pH (Narwal *et al*., 1999; Basta *et al.*, 2005; Antoniadis *et al.*, 2008). It is important to take in account that the bioavailability of the metal is determined by several factors including mineralogy, content of matter, capacity of cation exchange and temperature of the soil, additionally, the metal fraction and its location in the soil system (Takáč *et al.*, 2009; Nederlof *et al.*, 1993). For this problem, there are different methods for improvement of contaminated soils like removal of contaminated soil, adding uncontaminated soil as a

topdressing, use of a soil conditioner depending on the heavy metal on the soil and, improvement of water management (Iwata *et al.*, 1994).

In Latin America, Colombia is the third country with the highest water resources, climate diversity and annual precipitation rates, characteristics that favor its role in food production. Therefore, The United Nations Food and Agriculture Organization (FAO) affirms that Colombia has great potential to be a pantry of the world. However, in order to supply the world population, food production will have to be increased by 70% by the year 2050 (FAO, 2018). In this context, to ensure food security, regulatory measures should be implemented for the use of chemical products in agricultural activities due it is a problem in areas of great potential and the production, besides being not sustainable and causes serious environmental damage. Particularly, it is probably that in southern zone of the ID of Repelón, the low populations of nitrogen fixing bacterial reported by Martínez-Mera *et al.* (2007) are associated with the soil properties and high concentrations of heavy metals compared with the entire study zona. Puga *et al*. (2006) affirm that the problems with heavy metals including loss of diversity, decrease of the biological potential of the soil, zero agricultural productivity, and effects on public health as chronic-degenerative diseases in people, among others.

Conclusions

In the studied agricultural soils, no organochlorine or organophosphorus pesticides were detected, possibly the scarce agricultural activity at the time of sampling, the variability of the climatic conditions of the area that could favor the natural attenuation of the pesticides and the characteristics of the soils and/or compliance with agricultural regulations. In contrast, the concentration of HM in the soils varied as follows Zn>Cr>Ni>Pb> Hg>Cd. The pollution indicators showed contamination in the study area, that may be related to anthropogenic activities (quarries activities, agricultural practices and, urbanization). And the natural phenomenon such as high rainfall and flooding of soils. Likewise, the canonical analysis showed in the first correlation a positive correlation between silt and clay with Cd, and a negative correlation between Ni and Hg, with silt and clay. In the second correlation showed a positive relationship was observed between silt and Hg, Zn and Cr, and a negative relationship between CEC with Hg, Zn and Cr. In the third correlation a high negative correlation was found between the pH and the Cd, Cr and Pb concentrations. Soil properties like OM, CEC, silt and, clay contributes to HM retention by decreasing tis bioavaility.

This research provides knowledge on the environmental health of the agricultural soils of the Repelón ID. Information that can serve as a tool for defining strategies for sustainable management for the implementation of monitoring plans to quantify the vulnerability of the ecosystem and potential risk to human health being that the population depends on these natural resources for the development of their economic activities.

Acknowledgments

The authors thanks to Tito Crissien, General Manager of Universidad de la Costa. His collaboration and support was invaluable for developing the research project INV.1106-01-001-11. Authors acknowledge and appreciate the participation of Jariff Garrido, Maria Licona and Karols Scaldaferro for their continued support during the sampling campaigns and laboratory analysis. And special thanks to the Laboratory of Toxicology and Environmental Management at the Universidad de Córdoba for the chemical analysis.

References

Ahumada, I., Mendoza, J., Ascar, L. 1999. Sequential extraction of heavy metals in soils irrigated with wastewater. Commun. Soil Sci. Plant Anal., 30, 1507-1519. https://doi.org/10.1080/00103629909370303

Alcaldía de Repelón-Atlántico. 2016. Información general. [Online] Available at: http://www.repelon-atlantico.gov.co/informacion_general.shtml#geografia [Accessed 15 March 2017].

Alcaldía de Repelón-Atlántico. 2017. Información general. [Online] Available at: http://www.repelonatlantico.gov.co/index.shtml?apc=gbxx2760911&sh_itm=98bd43dd20b101b40297b595248a69a4&add_disc=1. [Accessed 20 November 2017].

Alloway, B.J. 2013. Sources of heavy metals and metalloids in soils. In Heavy metals in soils (Alloway, B.J. Ed.). Trace Metals and Metalloids in Soils and their Bioavailability. Third edition. Springer, 11-50. http://doi.org/10.1007/987-94-007-4470-7

Andrades, M., Martínez, E. 2014. Fertilidad del suelo y parámetros que la definen. Tercera Edición. Universidad de la Rioja. Logroño: España, 34.

Angelova, V.R., Akova, V.I., Artinova, N.S., Ivanov, K.I. 2013. The effect of organic amendments on soil chemical characteristics. Bulg. J. Agric. Sci., 19, 5, 958-971.

Antoniadis, V., Robinson, J., Alloway, B. 2008. Effects of short-term pH fluctuations on cadmium, nickel, lead, and zinc availability to ryegrass in a sewage sludge-amended field. Chemosphere, 71, 759–764. http://doi:10.1016/j.chemosphere.2007.10.015

Basta, N.T., Ryan, J.A., Chaney, L. 2005. Trace element chemistry in residual-treated soil: key concepts and metal bioavailability. J. Environ. Qual, 34, 49–63.

Bautista-Cruz, A., Etchevers-Barra, J., Del Castillo, R. F., Gutiérrez, C. 2004. La calidad del suelo y sus indicadores. Ecosistemas, 13, 2, 90-97.

Belmonte-Serrato, F., Romero-Díaz, A., Moreno-Brotóns, J. 2010. Contaminación ambiental por estériles mineros en un espacio turístico en desarrollo, la sierra minera de Cartagena-La Unión (sureste de España). Cuad. Turismo, 25, 11-24.

Buchman, M.F. 2008. NOAA Screening Quick Reference Tables. [Online] Available at: 8310https://repository.library.noaa.gov/view/noaa/9327. [Accessed 16 November 2017].

Carrillo, R., Cajuste, L. 1995. Behavior of trace metals in soils of Hidalgo, México. J. Environ. Sci. Health. B., 30, 143-155.

Castellanos, J. 2016. Manual para la interpretación de análisis de suelo. [Online] Available at: http://www.fec-chiapas.com.mx/sistema/biblioteca_digital/guia-de-interpretacion-de-analisis-de-suelos-y-aguas-intagri-3.pdf/. [Accessed 28 October 2017].

Cornejo, J., Jamet, P. 2000. Pesticide/soil interactions: some current research methods. Institut National de la Recherche Agronomique (INRA). Paris, Francia.

Chen, X., Xia, X., Zhao, Y., Zhang P. 2010. Heavy metal concentrations in roadside soils and correlation with urban traffic in Beijing, China. J. Hazard. Mater., 181, 1-3, 640–646. https://doi.org/10.1016/j.jhazmat.2010.05.060

Climate, Data. 2017. National Center for Environmental Information. [Online] Available at: http://es.climate-data.org/location/50352/. [Accessed 1 March 2017].

CRA – CRADIQUE, Corporación Autónoma Regional del Atlántico y Corporación Autónoma del Canal del Dique. 2002. Ministerio del Medio Ambiente – Colombia. 243p.

CRA - Corporación Autónoma Regional del Atlántico. 2014. Diagnóstico inicial para el ordenamiento del embalse El Guájaro y la ciénaga de Luruaco. Barranquilla, Atlántico.

De Alba, S., Torri, D.; Borselli, L., Lindstrom, M. 2003. Degradación del suelo y modificación de los paisajes agrícolas por erosión mecánica (Tillage erosion). J. Soil Sci. 10, 3, 93-101.

FAO-United Nations Food and Agriculture Organization. 2018. World agriculture: towards 2015/2030. In: Food and agriculture in national and international environments. Available at http://www.fao.org/docrep/004/y3557s/y3557s07.htm. [Accessed 29 January 2019].

Galán-Huertos, E., Romero-Baena, A. 2008. Contaminación de suelos por metales pesados. Macla, 10, 48-60.

García, Y., Ramírez, W., Sánchez, S. 2012. Indicadores de la calidad de los suelos: una nueva manera de evaluar este recurso. Pastos y Forrajes, 35, 2, 125-138.

Gaw, S. K., Wilkins, A. L., Kim, N. D., Palmer, G. T., Robinson, P. 2006. Trace elements and ΣDDT concentrations in horticultural soils from the Tasman, Waikato and Auckland regions of New Zealand. Sci. Total Environ., 355, 1–3, 31-47 https://doi.org/10.1016/j.scitotenv.2005.02.020

Gilden R. C., Huffling K., B. Sattle. 2010. Pesticides and health risks. J. Obstet. Gynecol. Neonatal Nurs., 39, 1, 103-10. https://doi.org/10.1111/j.1552-6909.2009.01092.x

Gimeno-García, E., Andreu, V., Boluda, R. 1996. Incidence of heavy metals in the application of inorganic fertilizers to rice farming soils (Valencia, Spain). In: Rodriguez-Barrueco C. (Eds) Fertilizers and Environment. Dev. Plant Soil Sci. Springer, Dordrecht. https://doi.org/10.1007/978-94-009-1586-2_85

Guangwei, D., Xiaobing, L., Stephen, H., Jeffrey, N., Dula, A., Baoshan, X. 2006. Effect of cover crop management on soil organic matter. Geoderma, 130, 3-4, 229-239. https://doi.org/10.1016/j.geoderma.2005.01.019

Hakanson, L. 1980. An ecological risk index for aquatic pollution control. A sedimentological approach. Water Res., 14, 975–1001.

Hoskins, B. 2017. Interpreting soil test results for gardens and grounds. Maine Soil Testing Service. University of Maine. [Online] Available at https://extension.umaine.edu/gardening/manual/soils/interpretting-soil-tests/. [Accessed 19 November 2017].

ICA-Instituto Colombiano Agropecuario. [Online] Available at https://www.ica.gov.co/getdoc/b2e5ff99-bd80-45e8-aa7a-e55f0b5b42dc/PLAGUICIDAS-PROHIBIDOS.aspx. [Accessed 19 November 2017].

IDEAM- Instituto de Hidrología, Meteorología y Estudios Ambientales. 2017. Estudio Nacional del Agua. Bogotá a, D.C. 253p.

IGAC- Instituto Geográfico Agustín Codazzi. 2006. Métodos Analíticos del Laboratorio de Suelos Sexta edición. Bogotá: Imprenta Nacional de Colombia. 647p.

IGAC- Instituto Geográfico Agustín Codazzi. 2008. Estudio general de suelos y zonificación de tierras. Departamento del Atlántico. Bogotá: Imprenta Nacional de Colombia. 324p.

Iqbal, J., Saleem, M., Shah, M.H. 2016. Spatial distribution, environmental assessment and source identification of metals content in surface sediments of freshwater reservoir, Pakistan. Chem. Erde-Geochem, 76, 1, 171–177. http://doi.org/10.1016/j.chemer.2016.02.002

Iwata, S., Tabuchi, T., Warkentin, B.P. 1994. Soil-Water Interactions: Mechanisms Applications. Second edition. CRC, Press, New York. 464p.

Jaurixje, M., Torres, D., Mendoza, B., Henríquez, M., Contreras, J. 2013. Propiedades físicas y químicas del suelo y su relación con la actividad biológica bajo diferentes manejos en la zona de Quíbor, Estado Lara. Bioagro, 25, 1, 47-56.

Jiao, W., Chen, W., Chang, A.C., Page, A.L. 2012. Environmental risks of trace elements associated with long-term phosphate fertilizer applications: a review. Environ. Pollut., 168, 44-53. https://doi.org/10.1016/j.envpol.2012.03.052.

Jung, M.C. 2008. Heavy Metal concentrations in soils and factors affecting metal uptake by plants in the Vicinity of a Korean Cu-W Mine. Sensors (Basel), 8, 4, 2413–2423.

Kabata-Pendias, A. 2011. Trace elements in soils and plants. Fourth edition. CRC, Press, Boca Ratón, FL, USA., 505p.

Kashem, M.A., Singh, B.R. 2001. Metal availability in contaminated soils: I. Effects of flooding and organic matter on changes in Eh, pH and solubility of Cd, Ni and Zn. Nutr. Cycl. Agroecosys., 61, 3, 247-255. https://doi.org/10.1023/A:1013762204510

Kim, H.K., Jang, T.I., Kim, S.M., Park, S.W. 2015. Impact of domestic wastewater irrigation on heavy metal contamination in soil and vegetables. Environ. Earth Sci., 73, 5, 2377-2383. https://doi.org/10.1007/s12665-014-3581-2

Marković, M., Cupać, C., Đurović, R., Milinović, J., Kljajić, P. 2010. Assessment of heavy metal and pesticide levels in soil and plant products from agricultural area of Belgrade, Serbia. Arch. Environ. Contam. Toxicol., 58, 2, 341-51. http://doi.org/10.1007/s00244-009-9359

Marrugo-Negrete, J. Pinedo-Hernández, J., Díez, S. 2017. Assessment of heavy metal pollution, spatial distribution and origin in agricultural soils along the Sinú River Basin, Colombia. Environ. Res., 154, 380-388. http://dx.doi.org/10.1016/j.envres.2017.01.021.

Martínez, E., Fuente, J., Acevedo, E. 2008. Carbono orgánico y propiedades del suelo. J. Soil Sci. Plant Nutr., 8, 1, 68-96.

Martínez-Mera, E.A., Torregroza-Espinosa, A.C., Valencia-García, A., Rojas-Gerónimo, L. 2017. Distribution of nitrogen fixing bacterial isolates and its relationship to the physicochemical characteristics of southern agricultural soils of the Atlántico department, Colombia. Soil Environ., 36, 2, 174-181. http://doi.org/10.25252/SE/17/51202

Melo, V.F.; Novais, R.F.; Fontes, M.P.F.; Schaefer, C.E.G.R. 2000. Potassium and magnesium minerals from sand and silt fractions of different soils. Rev. Bras. Cienc. Solo., 24: 269-284. http://dx.doi.org/10.1590/S0100-06832000000200004

Mir-Mohammad, A., Mohammad-Lokman, A., Md.-Saiful, I., Md.-Zillur, R. 2016. Preliminary assessment of heavy metal in water and sediment of Karnaphuli River, Bangladesh. Environ. Nanotechnol. Monit. Manage., 5, 27-35. https://doi.org/10.1016/j.enmm.2016.01.002

Müller, G. 1981. Die Schwermetallbelastung der Sedimenten des Neckars und Seiner Nebenflu¨sse. Chem. Ztg., 6, 157– 64.

McKenzie, R. 2013. Soil Salinity - Causes and Management. [Online] Available at http://www.thecropsite.com/articles/1500/soil-salinity-causes-and-management/. [Accessed 12 December 2017].

Narváez, J., Palacio, J., Molina, F. 2012. Environmental persistence of pesticides and their ecotoxicity: A review of natural degradation processes. Gest. Ambiente, 15, 3, 27-38.

Narwal, R.P., Singh, B.R., Selbu, B. 1999. Association of Cd, Zn, Cu y Ni with components in naturally heavy metal rich soils studied by parallel and sequential extraction. Commun. Soil Sci. Plant Anal., 30, 1209-1230. https://doi.org/10.1080/00103629909370279

Nederlof, M.M., Van-Riemsdijk, W.H., De Haan, F.A.M. 1993. Effect of pH on the bioavailability of eetals in soils. In: Eijsackers, H.J.P. Hamers T. (Eds) Integrated Soil and Sediment Research: A Basis for Proper Protection. Soil and Environment, Springer, Dordrecht.

Novello, O.A., Quintero, C. E. 2009. Contenidos de fósforo total en suelos distrito Villa Eloisa (Santa Fé). Inf. Agronómicas, 41, 11-15.

NTC-Norma Técnica Colombiana - 5264. 2008. Calidad de Suelo, Determinación del pH. Bogotá, D. C, Colombia.

NTC-Norma Técnica Colombiana - 5889. 2011. Análisis de Suelos, Determinación del Nitrógeno Total. Bogotá, D. C, Colombia. 8p.

NTC-Norma Técnica Colombiana - 5268. 2014. Calidad de Suelo. Determinación de la capacidad de intercambio catiónico. Bogotá, D. C, Colombia. 8p.

Olsen, S.R., C.V. Cole, F.S. Watanabe., Dean, L.A. 1954. Estimation of available phosphorus in soils by extraction with sodium bicarbonate. U.S. Department of Agriculture Circ. 939.

Poot, A., Gillissen, F., Koelmans, A. 2007. Effects of flow regime and flooding on heavy metal availability in sediment and soil of a dynamic river system. Environ Pollut., 148, 779–787. http://dx.doi.org/10.1016/j.envpol.2007.01.045

Puga, S., Sosa, M., Lebgue, T., Quintana, C., Campos, A. 2006. Contaminación por metales pesados en el suelo provocado por la industria minera. Ecología Aplicada, 5, 1–2, 149-155.

Reichaman, M.S. 2002. The responses of plants to metals toxicity: A review focusing on copper, manganese and zinc. Australian Minerals and Energy Environment Foundation 54p. Melbourne, Australia.

Roqueme, J., Pinedo, J., Marrugo, J., Aparicio, A. 2014. Metales pesados en suelos agrícolas del valle medio y bajo del rio Sinú, departamento de Córdoba. Memorias del II Seminario de Ciencias Ambientales Sue-Caribe & VII Seminario Internacional de Gestión Ambiental, 2014. Universidad de Córdoba, Montería. Colombia.

Rueda-Saá, G., Rodríguez-Victoria, J.A., Madriñán-Molina, R. 2011. Methods for establishing baseline values for heavy metals in agricultural soils: Prospects for Colombia. AcAg., 60, 3, 203-218.

Ruíz-Cabarcas, A.C., Pabón-Caicedo, J.D. 2013. Efecto de los fenómenos del niño y la niña en la precipitación y su impacto en la producción agrícola del departamento del Atlántico, Colombia. Cuad. Geogr., 22, 2, 35–54.

Silva S.M., Correa F.J. 2009. Análisis de la contaminación del suelo: revisión de la normativa y posibilidades de regulación económica. Semestre Económico, 12, 23, 13-34

Silveira, M., Alleoni, L., Guilherme, L. 2003. Biosolids and heavy metals in soils. Sci. Agric. (Piracicaba, Braz.), 60, 4, 793–806. http://dx.doi.org/10.1590/S0103-90162003000400029

Simón, M., Peralta, N., Costa, J.L. 2013. Relación entre conductividad eléctrica aparente con propiedades del suelo y nutrientes. Cienc. Suelo, 31, 1, 45-55.

Suñer, L., Galantini, J., Rosell, R., Chamadoira. M. 2001. Cambios en el contenido de las formas de fósforo en suelos de la región semiárida pampeana cultivados con trigo. Revista Fac. Agron., 104, 2, 113-119.

Sparks, D. L. 2003. Environmental Soil Chemistry. Second edition. CA: Academic Press. San Diego, CA., 352.

Takáč, P., Szabová, T., Kozáková, Ľ., Benková, M. 2009. Heavy metals and their bioavailability from soils in the long-term polluted Central Spiš region of SR. Plant Soil Environ., 55, 4, 167-172

Tomlinson, D.C., Wilson, D.J., Harris, C.R., Jeffrey D.W. 1980. Problem in heavy metals in estuaries and the formation of pollution index. Helgol. Wiss. Meeres., 33, 1–4, 566–575.

Torregroza‐ Espinosa, A.C., Martínez‐Mera, E.A., Castañeda‐Valbuena, D., González Márquez, L.C., Torres‐Bejarano, F.M. 2018. Contamination level and spatial distribution of heavy metals in wáter and sediments of El Guájaro Reservoir, Colombia. Bull. Environ. Contam. Toxicol., 101, 61-67. https://doi.org/10.1007/s00128-018-2365-x

UPRA–Unidad de Planificación Rural Agropecuaria. 2013. Uso agrícola. En UPRA, Leyenda de usos agropecuarios del suelo. Bogotá: Imprenta Nacional.

US-EPA, United States-Environmental Protection Agency. 2007a. SW-846 Test Method 8081B: Organochlorine pesticides by gas chromatography. p 57.

US-EPA, United States-Environmental Protection Agency. 2007b. SW-846 Test Method 8141B: Organophosphorus pesticides by gas chromatography. p. 57.

US-EPA, United States-Environmental Protection Agency. 2007c. SW-846 Test Method 3051A: Microwave assisted acid digestion of sediments, sludges, soils, and oils. p. 30.

US-EPA, United States-Environmental Protection Agency. 2007d. SW-846 Test Method 7471B: Mercury in Solid or Semisolid Waste (Manual Cold-Vapor Technique). p. 11.

Uzcátegui, J., Araujo, Y., Mendoza, L. 2011. Residuos de plaguicidas organoclorados y su relación con paramétros físicoquímicos en suelos del municipio Pueblo Llano, Estado Mérida. Bioagro, 23, 2, 115-120.

Vallejo, P., Vásquez, L., Correa, I., Bernal, G., Alcántara, J., Palacio, J. 2016. Impact of terrestrial mining and intensive agriculture in pollution of estuarine surface sediments: spatial distribution of trace metals in the Gulf of Urabá, Colombia. Mar. Pollut. Bull., 111(1-2), 311-320. https://doi.org/10.1016/j.marpolbul.2016.06.093

Wu, G., Wu, J.Y., Shao, H.B. 2012. Hazardous heavy metal distribution in Dahuofang Catchment, Fushun, Liaoning, an important industry city in China: A Case Study. Clean - Soil, Air, Water, 40, 12, 1372–1375. http://doi.org/ 10.1002/clen.201000589

Yacomelo, M. 2014. Riesgo toxicológico en personas expuestas, a suelos y vegetales, con posibles concentraciones de metales pesados, en el sur del Atlántico, Colombia. Tesis Maestría. Universidad Nacional de Colombia.

Yap, B.P., Sim, C.H. 2011. Comparisons of various types of normality tests. J. Stat. Comput. Sim., 81, 12, 2141-2155. https://doi.org/10.1080/00949655.2010.520163.ord

Weather Online. 2018. Available at https://www.worldweatheronline.com/lang/en-au/repelon-weather-averages/atlantico/co.aspx [Accessed 04 March 2018].

Wuana, R.A., Okieimen, F.E. 2011. Heavy metals in contaminated soils: A review of sources, chemistry, risks and best available strategies for remediation. Int. Sch. Res. Notices. Article ID 402647, 20 pages.

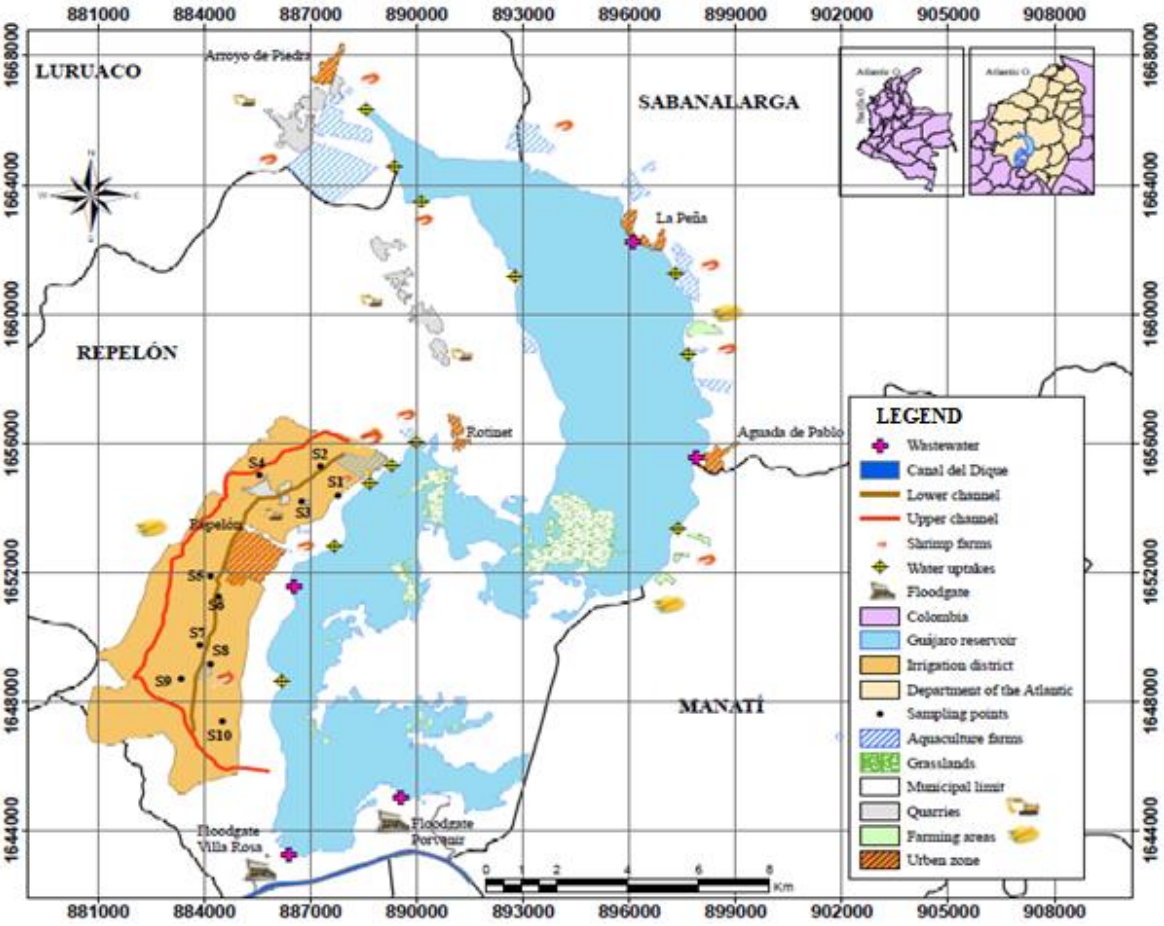

Figure 1. General location of the ID of Repelón.

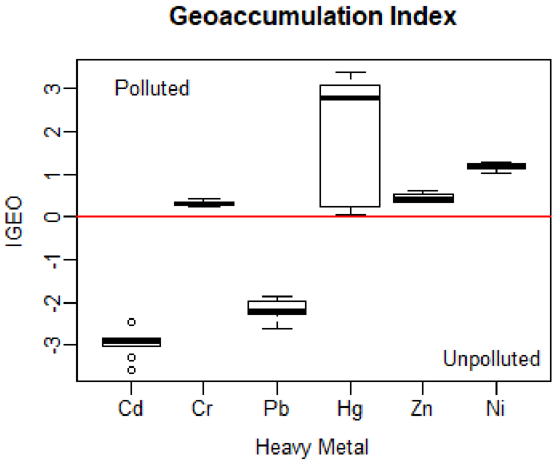

Figure 2. Geoaccumulation index (Igeo) of the metals analyzed in the agricultural soils of the Repelón ID.

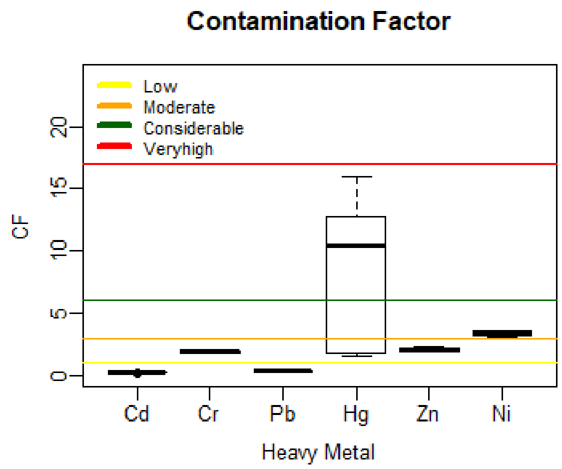

Figure 3. Contamination Factor (CF) of heavy metals found in agricultural soils of the Repelón ID.

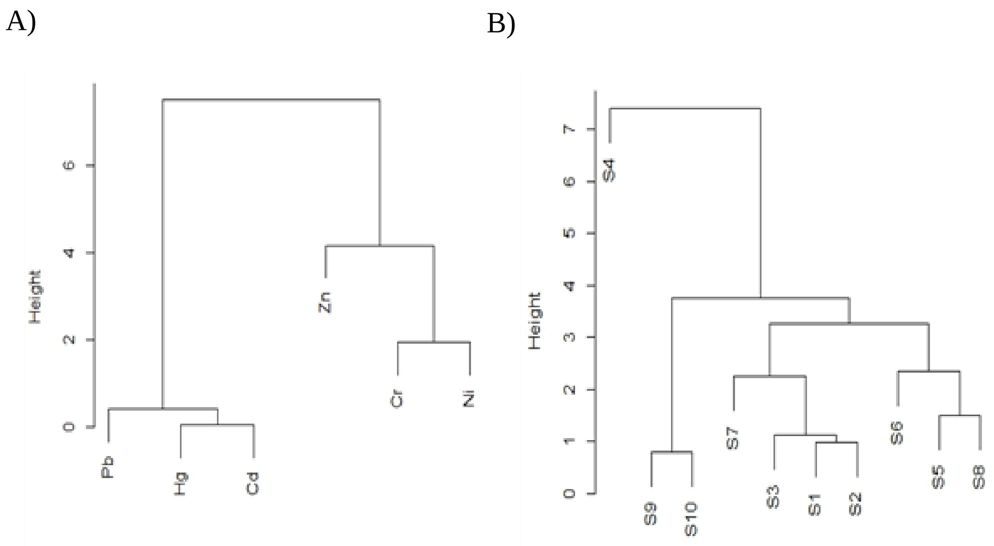

Figure 4. Dendrogram obtained by hierarchical clustering analysis for (A) the heavy metals, and (B) the sampling sites.