

# 1 Diatom inferred aquatic impacts of the mid-Holocene eruption of Mount 2 Mazama, Oregon

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11

## 12 ABSTRACT

13

14 High-resolution diatom stratigraphies from mid-Holocene sediments taken from fringe and  
15 central locations in Moss Lake, a small lake in the foothills of the Cascade Range,  
16 Washington, have been analysed to investigate the impacts (and duration) of tephra  
17 deposition on the aquatic ecosystem. Up to 50 mm of tephra was deposited from the  
18 climactic eruption of Mount Mazama 7958-7795 cal yr BP, with coincident changes in the  
19 aquatic ecosystem. The diatom response from both cores indicates a change in habitat type  
20 following blanket tephra deposition, with a decline in tycho planktonic *Fragilaria brevistriata*  
21 and *Staurosira venter* and epiphytic diatom taxa indicating a reduction in aquatic macrophyte  
22 abundance. Additionally, the central core shows an increase in tycho planktonic *Aulacoseira*  
23 taxa, interpreted as a response to increased silica availability following tephra deposition.  
24 However, partial redundancy analysis provides only limited evidence of direct effects from

25 the tephra deposition, and only from the central core, but significant effects from underlying  
26 environmental changes associated with climatic and lake development processes. The  
27 analyses highlight the importance of duplicate analyses (fringe and central cores) and  
28 vigorous statistical analyses for the robust evaluation of aquatic ecosystem change.

29 Keywords: Tephra impact, Diatoms, Mazama, Redundancy analysis, Holocene, Volcano

30

## 31 INTRODUCTION

32

33 Volcanic eruptions may have major impacts on ecosystems, and although the proximal  
34 impacts are generally well understood, the distal impacts of distal tephra deposition are not  
35 well defined (see Payne and Egan, 2017 for a review). The ‘fine’ to ‘very fine’ (between <30  
36 and 100  $\mu\text{m}$ ) class size of ash is of particular importance in this context as they have the  
37 longest atmospheric residence times, travel the furthest distance and carry the most toxic  
38 volatiles (Payne and Blackford, 2008), which makes them particularly hazardous to the  
39 environment (Rose and Durant, 2009). The impact of volcanic eruptions on climate and  
40 ecosystems can be detrimental. This can be through direct climatic perturbations (Mass and  
41 Portman, 1989; McCormick *et al.*, 1995; Zielinski, 2000; Stoffel *et al.*, 2015), though such  
42 effects usually last only 1-3 years (McCormick *et al.*, 1995; Zdanowicz *et al.*, 1999). It is the  
43 effects of tephra deposition on different receiving environments that may have longer-term,  
44 decadal to centennial (Barker *et al.*, 2000; Telford *et al.*, 2004; Blackford *et al.*, 2014; Egan *et*  
45 *al.*, 2016) and even millennial (Bradbury *et al.*, 2004) effects and are more ambiguous in  
46 nature. Focussing on the aquatic effects of tephra deposition, this paper aims to enhance our  
47 understanding of tephra impacts.

48 Egan *et al.*, (2016) conducted a study on the terrestrial impacts of the tephra deposited by the  
49 Plinian eruption of Mount Mazama, Cascade Range, 7682-7584 cal yr BP (95.4% probability  
50 range) (Egan *et al.*, 2015). Mount Mazama ejected nearly 50 km<sup>3</sup> of rhyodacitic magma into  
51 the atmosphere (ten times as much as the 1980 eruption of Mount St Helens), and deposited  
52 ash over an area of approximately 1.7x10<sup>6</sup> km<sup>2</sup> (Zdanowicz *et al.*, 1999) in a predominantly  
53 north-easterly direction (Figure 1). Egan *et al.*, (2016) reported a significant local impact on  
54 the terrestrial environment surrounding Moss Lake, Washington, with decreases in the open  
55 habitat vegetation. Building on the potential impacts from the Mazama tephra deposit here,  
56 we are now primarily interested in the nature and duration of the aquatic impacts, which Egan  
57 *et al.*, (2016) did not address. Egan *et al.*, (2016) highlighted the importance of multiple cores  
58 as their analyses revealed a local terrestrial impact but no regional impact. Here we assess  
59 both shallow and deep cores from Moss Lake through the use of high resolution diatom  
60 analyses to determine the physiochemical effects of tephra deposition on the aquatic  
61 environment.

62 Tephra can impact aquatic systems physically, biologically and chemically. The direct input  
63 of tephra may instantly kill aquatic life as tephra may get stuck in fish gills and can be toxic  
64 to aquatic organisms (Ayris and Delmelle, 2012; Lallement *et al.*, 2016). Physically tephra  
65 can make the water more turbid (Lallement *et al.*, 2016), reduce light infiltration (Abella,  
66 1988), alter drainage patterns and water flow (Lallement *et al.*, 2016) and alter habitat  
67 availability with the death of aquatic plants reducing epiphytic species and newly deposited  
68 tephra allowing epipelagic algae to colonise (Telford *et al.*, 2004). Such physical changes can  
69 have implications for the biology as fish and algae can be displaced as their habitat is  
70 destroyed, and the changes in light penetration can reduce photosynthetic activity thus  
71 altering oxygen levels in the water. Chemically the aquatic system can be changed by  
72 increased concentrations of heavy metals (Power *et al.*, 2011) or by decreasing the pH

73 through the influx of associated sulphuric acid (Birks and Lotter, 1994). Silica is also added  
74 to the aquatic system. Whilst this is an instant input, Telford *et al.*, (2004) state that it  
75 dissolves slowly, and even when the tephra is deposited onto the lake bed, the silica will still  
76 be released, which provides a steady influx of silica to the aquatic system. Some diatom  
77 species are likely to be competitively advantaged and response positively to the increased  
78 silica input (Telford *et al.*, 2004). An additional effect of tephra deposition in lakes is that a  
79 thick tephra layer creates an impermeable barrier over the lake's sediment, preventing the  
80 regeneration of nutrients such as phosphorus (Barker *et al.*, 2000, 2003). Tephra presents a  
81 physical barrier to the transport of phosphorus into the water column by preventing re-  
82 suspension of the sediment by bioturbation and wave action, as well as a barrier to  
83 phosphorus diffusion depending on its thickness. Any changes in the chemical status of the  
84 lake would impact on the biological and thus physical characteristics of the aquatic system, as  
85 these three components are interconnected.

86 A recent study has assessed the aquatic impacts of tephra fallout from the rift zone eruption  
87 of the Puyehue Cordon-Caulle volcanic complex, Patagonia in 2011, and reported habitat  
88 loss, changes in morphology of the main channels, increases of turbidity and a sharp decline  
89 in salmonid fish densities (Lallement *et al.*, 2016). These impacts persisted for 30 months  
90 after the initial eruption, with some evidence that recovery is underway but uncertainties exist  
91 as to whether the channels and their fish assemblages will ever return to the pre-impact  
92 conditions. Continuous monitoring in active volcanic terrains like this example is rare and  
93 few have the decadal durations required to record longer-term trends, thus much less is  
94 known about the long-term impacts on aquatic ecosystems.

95 Palaeoenvironmental records have been used in several studies to successfully infer long-  
96 term impacts of volcanic eruptions and subsequent recovery. Diatoms are frequently used as a

97 proxy for tephra impacts in lakes as these algae are affected first due to their high sensitivity  
98 and rapid responses, and their preservation means that longer term impacts can be assessed  
99 (Abella, 1988; Birks and Lotter, 1994; Barker *et al.*, 2000; Colman *et al.*, 2004; Telford *et al.*,  
100 2004). They are particularly useful for determining chemical and physical changes in the  
101 aquatic system.

102 Six studies have assessed the aquatic impacts of tephra deposition from Mount Mazama  
103 (Blinman *et al.*, 1979; Abella, 1988; Hickman and Reasoner, 1994; Bradbury *et al.*, 2004;  
104 Stone, 2005). The studies reported varying impacts including: a decrease of pH (Bradbury *et*  
105 *al.*, 2004), increase of salinity (Heinrichs *et al.*, 1999; Bradbury *et al.*, 2004), a change in the  
106 Si:P ratio (Abella, 1988; Hickman and Reasoner, 1994) and changes in water chemistry that  
107 lasted nearly a millennium (Stone, 2005). Most of these lake systems are large in size and  
108 highly productive (Abella, 1988; Bradbury *et al.*, 2004; Stone, 2005), and therefore not  
109 typical or representative of the majority of lakes in the Cascade region which are smaller and  
110 nutrient poor (Hickman and Reasoner, 1994; Heinrichs *et al.*, 1999). Evaluations are  
111 therefore needed from such systems, which may be more sensitive to disturbance. This study  
112 provides an assessment of the distal impacts of tephra deposition on the aquatic ecosystem of  
113 a small, oligotrophic lake system within the foothills of the Cascades.

114

## 115 **STUDY AREA**

116

117 The Mazama tephra is of great stratigraphic importance as the climactic eruption of Mount  
118 Mazama was a high magnitude event producing the most significant tephra fall of the  
119 Holocene in North America and it has been identified at Moss Lake. Moss Lake (47°41.60N,  
120 121°50.81W – Datum used: WGS 84) is 500 km north of Crater Lake (the site of the Mount

121 Mazama eruptions) allowing distal impacts to be assessed. Moss Lake is located in a mixed  
122 conifer forest within the Tolt River basin in King County in the Cascade foothills at an  
123 elevation of 158 m (Figure 1). The lake has a diameter of approximately 200 m with a  
124 maximum depth of 4.5 m. Moss Lake resides in a shallow basin within a broad fluted basal  
125 till plain that was deposited during the Vashon Stade (~18,000-16,500 cal yr BP at the site;  
126 Porter and Swanson, 1998). This till sheet overlies glaciomarine drift and outwash deposits  
127 (Dragovich *et al.*, 2002).

128 During the time of the deposition event Moss Lake included an extensive shallow water  
129 system, and today the lake is surrounded by a reed swamp (see Figure 1, inset in top left)  
130 underlain by lake sediments. This lake setting is ideal for the study with stratigraphies from  
131 both a core from the fringing reed swamp, representative of the Holocene shallow water  
132 system and potentially dominated by benthic taxa, as well as a deep water core from the  
133 centre of the lake, potentially with higher representation of planktonic and tycho planktonic  
134 taxa. Moss Lake is a freshwater, oligotrophic system with a weakly acidic-neutral pH of 6  
135 and low conductivity of 14-22  $\mu\text{S cm}^{-1}$ . Water chemistry analyses indicate the lake has a low  
136 concentration of calcium suggesting that it has a low buffering capacity and may be sensitive  
137 to acid deposition from volcanic events (for present day water chemistry see Supplementary  
138 Table 1). This study will add further detail to current knowledge regarding impacts of tephra  
139 deposition, and being in such close proximity to other major volcanoes such as Glacier Peak,  
140 Mount St Helens and Mount Rainier it is essential that as much research is done in this area  
141 as possible.

142

## 143 MATERIALS AND METHODS

144

145 **Core collection**

146

147 To elucidate true volcanic impacts cores were taken from the lake fringe (MLF) and a central  
148 part of the lake (MLC) representative of a shallow water and deep water lake system. MLC  
149 was collected from the deepest point of Moss lake, determined with an echo sounder, using a  
150 modified Livingstone corer. A second core was extracted from the fringe (MLF) using a  
151 Russian corer. The cores were placed in guttering, wrapped in cling film and stored in the  
152 cold room (dark, 2-4°C) at The University of Manchester. As the focus of this paper is the  
153 impact of tephra deposition, analyses concentrate on sediments above and below the Mazama  
154 tephra deposit, not the whole Holocene record.

155

156 **Stratigraphic analyses and radiocarbon dating**

157

158 Data reported here for the stratigraphy and chronology are taken from Egan *et al.*, (2016)  
159 where further details of the methods can be found. To summarise, we measured the organic  
160 matter content, carbonates, particle size and magnetic susceptibility. For the analyses here,  
161 we report the data from the samples taken every 5 mm above, below, and through the tephra  
162 layer (magnetic susceptibility was measured every 10 mm). Particle size analysis was carried  
163 out in order to assist with the determination of the tephra layer boundary in MLF, as it was  
164 not distinct. Samples were taken every 10 mm (every 5 mm through the tephra layer).

165 The Mazama tephra layer has previously been geochemically identified on the JEOL-  
166 JXA8600 electron microprobe at the Research Laboratory for Archaeology and the History of  
167 Art, University of Oxford (Egan *et al.*, 2016).

168

169 AMS radiocarbon dates were obtained for both cores as discussed in the previous work in  
170 Egan *et al.*, (2016). Radiocarbon dating was carried out on bulk sediments as there were no  
171 identifiable macrofossils or macrocharcoal fragments. The low sedimentary carbonate content  
172 indicates that hard water reservoir effects are unlikely. Originally, eight radiocarbon dates  
173 were obtained for MLC (Egan *et al.*, 2016), here we present four of those ages that focus  
174 directly on the Mount Mazama deposit. Radiocarbon dates were calibrated to calendar years  
175 (cal yr BP) using OxCal v.4.2.4 (Bronk Ramsey, 2014), and the IntCal13 calibration curve  
176 (Reimer *et al.*, 2013). A full chronology was determined with an age-depth model. We used a  
177 *P\_sequence* deposition model in OxCal v.4.2.4. To account for the 40 mm thick tephra layer,  
178 representing instantaneous deposition an “event free depth scale” was included (Staff *et al.*,  
179 2011). Three radiocarbon ages are reported for MLF, however an age reversal was present so  
180 an accurate chronology could not be determined.

181

## 182 **Diatom analysis**

183

184 High resolution samples (1 mm contiguous samples for MLF and 5 mm contiguous samples  
185 for MLC) were taken before and after the tephra layers. The age-depth model for MLC  
186 suggests these samples represent approximately 10-20 years. The high resolution sampling  
187 was done by slicing the sediment with a scalpel and avoiding areas of tephra penetration  
188 outside of the primary tephra layer. Diatom preparation followed the standard procedure by  
189 Battarbee (1986), and followed Renberg's (1990) recommendation of bulk preparation using  
190 a water-bath. Approximately 0.03 g of dry sediment was digested in 5 ml of 30% hydrogen



191 peroxide, 1-2 drops of hydrochloric acid were added to eliminate any remaining hydrogen  
192 peroxide and carbonates. Samples were washed several times and weak ammonia was added  
193 on the final wash to keep clays in suspension and to prevent diatom clumping. Microspheres  
194 were added to determine diatom concentration (Battarbee and Kneen, 1982). The  
195 concentration of microspheres added was 2 ml of  $5.01 \times 10^6$  per 0.01 g dry weight of sediment.  
196 Samples were mounted on the microscope slide using Naphrax<sup>®</sup> and were identified and  
197 counted at 1000x magnification. Identification was through the website “Diatoms of the  
198 United States” (Spaulding, 2014) and identification keys (Krammer and Lange-Bertalot,  
199 1991, 1999a, 1999b). At least 300 diatom frustules were counted.

200 Diatom diagrams presented here show the percentages of total frustules. The summary  
201 diagram is based on habitat preference determined primarily from “Diatoms of the United  
202 States” (Spaulding, 2014) and Kelly *et al.*, (2005). Diatom zonation was used not only to  
203 assist with qualitative analyses, but also as a quantitative tool, as the zones determined  
204 represent significant changes in the assemblage. To determine statistically significant changes  
205 optimal splitting by information content was used (Bennett, 1996), and the number of  
206 significant zones was determined through the use of the Broken-Stick model (Bennett, 1996).  
207 Diatom diagrams and the zonation were created using Psimpoll v.4.27 (Bennett, 2007).

208

### 209 **Ordination and associated significance tests**

210

211 Ordination was used to test for significant changes in the diatom record following the  
212 deposition of tephra, evaluating the significance of the impact of each tephra relative to and  
213 independently from additional environmental variables chosen to account for underlying

214 environmental trends. Detrended Correspondence Analysis (DCA) (Hill and Gauch, 1980)  
215 was used initially to estimate the length of the gradients in the biostratigraphical data sets (in  
216 standard deviation units). The diatom assemblages have short gradients (<1.7 SD), and  
217 consequently linear ordination methods were employed (Leps and Šmilauer, 2014). Principal  
218 Component Analysis (PCA) (Orloci, 1966) was then used to describe the relationships  
219 between different diatom species and samples.

220 Partial Redundancy Analysis (RDA) (ter Braak and Prentice, 1988; Rao, 1964), a constrained  
221 form of PCA, was used to determine how much of the variation is explained by the  
222 environmental variables and their significance. Log transformation and double centring of the  
223 samples and environmental variables were used to allow for the closed compositional  
224 disposition of the data. In order to test the significance of each environmental variable  
225 independent from the other two co-variables, timeseries restricted Monte Carlo permutation  
226 tests used for stratigraphically ordered data (ter Braak and Šmilauer, 2012) were completed  
227 with 999 permutations. The significance test compares eigenvalues for the first RDA axes of  
228 the diatom assemblages. The statistical program used was Canoco v5 (ter Braak and  
229 Šmilauer, 2012).

230 The influence of three independent environmental variables (tephra, LOI and core depth) on  
231 the diatom data was evaluated using direct ordination (partial RDA). Observed changes in the  
232 diatom assemblages around the time of volcanic events may have been a response to tephra  
233 deposition. This effect is modelled as an exponential decay function through time (Lotter and  
234 Birks, 1993; Birks and Lotter, 1994; Barker *et al.*, 2000; Lotter and Anderson, 2012;  
235 Blackford *et al.*, 2014). Prior to deposition of tephra the tephra explanatory variable was  
236 given a value of 0. At the time of tephra deposition, the value for ash was given an arbitrary  
237 value of 100, and after deposition the value of ash decreased exponentially  $x^{-\alpha t}$ , where  $\alpha$  is the  
238 decay coefficient and  $t$  is sample time ( $f$ = depth) since tephra deposition. Three models

239 (different decay coefficients) were used for the diatom assemblage from MLC to reflect  
240 different potential recovery times. Model 1 had a decay coefficient of 0.8 to reflect the  
241 longest recovery time of approximately 500 years, with most recovery having happened  
242 within approximately 200-250 years, model 2 had a decay coefficient of 0.5 to reflect  
243 medium duration or recovery of approximately 200 years, with most recovery having  
244 happened within approximately 100 years, and model 3 had a decay coefficient of 0.1 to  
245 reflect the shortest recovery time of approximately 80 years, with most recovery having  
246 happened within approximately 20 years. An ongoing study in an alpine lake in Washington  
247 has found strong evidence that tephra from Mount Mazama exerted significant influence on  
248 sedimentation dynamics for up to 500 years post deposition (Wershow, pers. comm.), thus  
249 the timeframes suggested by the models is realistic. For MLC all models were used as there  
250 was variation within the results. However, for MLF there was little difference in the results,  
251 so the decay coefficient used for this assemblage was 0.5, model 2. LOI was the second  
252 environmental variable representing the inflow of exogenic mineral materials, which would  
253 be associated with low organic matter content, and local environmental change. LOI was  
254 corrected for tephra by interpolating values for the samples containing tephra, so there was no  
255 influence of the tephra itself on this variable. The third variable was depth, as a surrogate for  
256 age to represent directional change during the period of tephra deposition associated with  
257 climate change or successional processes.

258

## 259 **RESULTS**

260

### 261 **Stratigraphy**

262 A wider stratigraphy for MLC is reported in Egan *et al.*, (2016), Figure 2 displays the  
263 stratigraphy of MLC around the time of Mazama tephra deposition (MLC-T324), the focus  
264 here. Figure 2 shows a shift from silty gyttja in sediment unit MLCs-1 to organic peaty silt in  
265 MLCs-2. LOI decreases and drops to 5% upon the deposition of the Mazama tephra. Organic  
266 matter content increases steadily after this in MLCs-2. Magnetic susceptibility increases upon  
267 tephra deposition from  $0.2 \times 10^{-7}$  SI to  $59 \times 10^{-7}$  SI. Carbonate content values are consistently  
268 below 0.01%. Figure 3 illustrates the MLF stratigraphy (again, a wider stratigraphy is  
269 presented in Egan *et al.*, (2016)). Particle size analysis was used to determine the boundary of  
270 tephra deposition and shows a peak in coarse and fine sand between 158-153 cm, indicative  
271 of the tephra boundary. At the base of sediment unit MLFs-1 organic sandy silts dominate  
272 with low LOI (~25%) and magnetic susceptibility values of  $3.5 \times 10^{-6}$  SI. Within the Mazama  
273 tephra deposit (MLF-T158) LOI further decreases to ~17% and magnetic susceptibility peaks  
274 to  $94 \times 10^{-6}$  SI. In MLFs-2 silty peats develop with an increasing LOI from ~20% to ~80%  
275 and generally low magnetic susceptibility of  $1-7 \times 10^{-6}$  SI. A silt unit is present from 146-132  
276 cm. There is a brief increase of magnetic susceptibility (to  $45 \times 10^{-6}$  SI) and particle size at  
277 around 120 cm where there is a coarse sand deposit. Carbonate content values are  
278 consistently below 0.03%.

279

## 280 **Radiocarbon**

281

282 The MLC sediment record extends back to the late Pleistocene, 16,294-12,789 cal yr BP,  
283 reported previously in Egan *et al.*, (2016) and is well constrained. However, the focus here is  
284 on the aquatic impact of the Mazama tephra deposit. Thus the record presented here spans the  
285 time period ~8400 cal yr BP to ~7100 cal yr BP (Table 1, Figure 4). The three radiocarbon

286 dates for MLF demonstrated an age reversal in the top two samples and was confirmed by re-  
287 analysis of the samples (Supplementary Table 3). The dates therefore cannot be used in the  
288 analyses, but are provided to demonstrate that MLF-T158 is within the right time period as  
289 the sediments below the tephra have a calibrated age range of 7958-7795 cal yr BP (95.4%  
290 probability range). Therefore, further up the core within the tephra layer the age is likely to be  
291 younger and within the previously published age range of 7682-7584 cal yr BP (95.4%  
292 probability range) (Egan *et al.*, 2015).

293

## 294 **Diatoms**

295

296 MLC has high proportions of planktonic *Discostella pseudostelligera* (up to 60%) and low  
297 proportions of epipelagic (~10%) and epiphytic (<5%) taxa. *Discostella pseudostelligera* is  
298 dominant throughout the assemblage. Figure 5 displays the diatom assemblage for MLC and  
299 Table 2 summarises the main changes in the diatom assemblage around the time of Mazama  
300 tephra deposition. Three zones have been identified. The first zone (C1) includes the pre-  
301 tephra assemblage and the tephra deposit itself. The transition into the second zone (C2)  
302 starts above the tephra deposit at 320.8 cm, the third zone (C3) starts at 313.7 cm. The  
303 zonation suggests tephra deposition from Mount Mazama caused marked change in the  
304 aquatic environment with clear differences in the assemblages before and after tephra  
305 deposition. In zone C1 *Discostella pseudostelligera* increasingly dominates from 40% to  
306 60%, and tycho planktonic species *Fragilaria brevistriata* and *Staurosira venter* decrease  
307 from 15% to <5% and 25% to <5% respectively. When tephra is deposited *Aulacoseira* taxa  
308 increase by a small percentage (<5%) and *Fragilaria brevistriata* and *Staurosira venter*  
309 decrease further to <5%. In zone C2 *Discostella pseudostelligera* still dominates but

310 decreases from 60% to 40%, whilst tychoplanktonic taxa increase overall from 30% to 45%.  
311 Zone C3 consists of a higher abundance of *Discostella pseudostelligera* (up to 70%),  
312 *Aulacoseira* taxa (>20%), *Fragilaria brevistriata* (up to 20%) and *Staurosira venter* (up to  
313 20%) than in zone C2.

314 MLF (Figure 6, Table 2) has a very different assemblage to MLC, in particular lower  
315 proportions of planktonic *Discostella pseudostelligera* (<5%) and higher proportions of  
316 epipellic (up to 50%) and epiphytic (up to 50%) taxa. The assemblage is dominated by  
317 epiphytic taxa throughout most of the profile, specifically *Eunotia soleirolii*, *Encyonema*  
318 *mesianum* and *Gomphonema gracile*. Tychoplanktonic taxa, specifically *Aulacoseira lirata*,  
319 have a high abundance (up to 55%) before the tephra deposition event, and epipellic taxa are  
320 high in abundance (up to 50%) after the tephra event, notably *Brachysira brebissonii*  
321 increasing from <5% to 20%. Two zones have been identified, and the split is during the time  
322 of Mazama tephra deposition, suggesting there was a marked assemblage change. Before  
323 tephra deposition in zone F1 epiphytic taxa dominate (up to 40%), in particular *Gomphonema*  
324 *gracile* and *Eunotia soleirolii* along with tychoplanktonic *Aulacoseira lirata*, *Aulacoseira*  
325 *alpigena*, *Fragilaria brevistriata* and *Staurosira venter*. Epipellic taxa increase after tephra  
326 deposition from 20% to 50%. In zone F2 epipellic species, especially *Brachysira brebissonii*  
327 dominate. Tychoplanktonic and epiphytic taxa are in lower abundance than in zone F1  
328 decreasing from up to 50% to 10% and 50% to 40% respectively.

329

### 330 **Ordination and significance tests (PCA and partial RDA)**

331

332 *Principal Components Analysis*

333 The gradients in both data sets for MLF and MLC had lengths of 1.1 and 0.7 SD units  
334 respectively. These short gradient lengths show that there was restricted turn-over in the  
335 diatom data as a standard deviation unit length of 4 would be indicative of complete species  
336 turn-over (Lepš and Smilauer, 2014).

337 Moss Lake Central (MLC) has two important PCA gradients (Figure 5); PCA axis 1 accounts  
338 for 26% of the variance and PCA axis 2 explains a further 21% of the variance. PCA axis 1 is  
339 associated with short term variations in tycho planktonic taxa (Figure 5) with a shift from  
340 positive to negative sample scores in zone C1, until Mazama tephra deposition in zone C2  
341 where sample scores increase and become positive. In zone C3 sample scores return to values  
342 observed in zone C1. Positive sample scores are driven by *Aulacoseira crassipunctata* and  
343 *Nitzschia palea*. Negative sample scores are driven by *Staurosira venter* and *Fragilaria*  
344 *brevistriata*.

345 PCA axis 2 is more strongly related to the diatom response to tephra deposition as there is a  
346 clear coherence between change in PCA axis 2 and the tephra deposition, specifically the  
347 responses of *Discostella pseudostelligera* and *Aulacoseira* taxa. Sample scores are weakly  
348 negative in zone C1, and become positive around the time of tephra deposition. In zone C2  
349 sample scores are variable but positive then decrease towards the top of the zone, and  
350 increases in zone C3 where sample scores fluctuate again. Positive sample scores are  
351 dominated by *Aulacoseira lirata* and *Aulacoseira alpigena*. The negative sample scores are  
352 dominated by *Discostella pseudostelligera* and *Tabellaria flocculosa*.

353 For MLF PCA axis 1 accounts for 50.39% of the variance and represents the dominant  
354 gradient in the diatom data (Figure 6). PCA axis 2 accounts for only a further 8% of the  
355 variance. Sample scores of PCA axis 1 are strong and positive in zone F1, then become  
356 weakly negative in zone F2 around the time of Mazama tephra deposition. After tephra

357 deposition sample scores show a steady, increasing trend reverting to scores similar to those  
358 of the pre-tephra assemblage but not fully back to baseline conditions. Positive sample scores  
359 are driven by *Aulacoseira alpigena*, *Staurosira venter*, and *Aulacoseira lirata*. The negative  
360 sample scores are driven by *Brachysira brebissonii*, *Frustulia rhomboides*, *Craticula*  
361 *halophila* and *Navicula bremensis*.

362

### 363 *Partial Redundancy Analysis*

364 For MLC partial RDA analyses (Table 3) revealed tephra to have a significant unique effect  
365 on the diatom assemblages in model 1, but not in models 2 and 3. Depth (directional change)  
366 had a significant unique effect in all analyses. LOI also had a significant unique effect except  
367 in MLC model 1 (Table 3).

368 For MLF the second model was applied to the dataset. In this model tephra was not  
369 significant, but depth and LOI had significant unique effects explaining 37.7% and 15.4% of  
370 the variance, respectively.

371 For MLC, the first model used an exponential decay rate of 0.8, assuming a 500 year  
372 recovery period with most recovery having happened within 200-250 years. This model  
373 reported tephra to have a significant unique effect on the diatom assemblage explaining  
374 11.2% of the variance. Depth also had a significant unique effect explaining a further 10.6%  
375 of the variance. The second model used an exponential decay rate of 0.5, assuming a 200 year  
376 recovery period, with most recovery having happened within 100 years. In this model tephra  
377 was not significant but depth and LOI indicated significant unique effects explaining 12.7%  
378 and 9.1% of the variance, respectively. The third model assumes a recovery period of 80  
379 years, with most recovery having happened within 20 years, through the application of a



380 decay rate of 0.1, and gave similar results to model 2 with tephra having no significant unique  
381 effect on the diatom assemblage but depth and LOI having a significant unique effect,  
382 explaining 12.6% and 11.3% of the variance, respectively (Table 3).

383

## 384 **DISCUSSION**

385

386 Analysis of the diatoms from MLF and MLC clearly illustrates two very different  
387 assemblages, with a low proportion of planktonic taxa in MLF (Figure 6) compared to MLC  
388 (Figure 5) consistent with shallow and deeper lake water systems respectively at the time of  
389 the climactic eruption of Mount Mazama. The response to tephra deposition may therefore be  
390 expected to differ between the two locations. The species mix confirms that Moss Lake is an  
391 oligotrophic, low alkalinity system with high proportions of *Aulacoseira* taxa and *Brachysira*  
392 *brebissonii*, indicating sensitivity to impacts associated with acidification and/or changes in  
393 the nutrient status. The DCA gradients in both data sets for MLF and MLC had lengths of  
394 1.1 and 0.7 SD units respectively, which suggests limited turn-over within the diatom  
395 assemblage (Lepš and Smilauer, 2014), however, changes are observed around the time of  
396 tephra deposition with MLC displaying a decline of *Discostella pseudostelligera*, and  
397 increases of *Aulacoseira* species, and MLF displaying notable decreases of *Aulacoseira*  
398 species and epiphytic taxa and increases of epipellic taxa, in particular *Brachysira brebissonii*.  
399 The partial RDA analyses for MLF and model 2 and 3 for MLC revealed that tephra from  
400 Mount Mazama overall had no unique significant effect on the diatom assemblages.  
401 However, model 1 for MLC was the only model indicating a significant unique effect  
402 (11.2%) of tephra deposition independent of variation in depth or LOI. Importantly all of the  
403 other models for both MLF and MLC indicate that depth (directional change) had the most

404 significant unique effect on the diatom assemblage (MLF, 37.7%; MLC, 10.6-12.7%) with  
405 LOI also having a significant influence (MLF, 15.5%; MLC, 9.1-11.3%).

406 There is therefore evidence of a tephra effect from the diatom analyses, supported by model 1  
407 of the partial RDA for MLC. However, the evidence for this tephra effect is not consistent  
408 between the two cores, or between the different models from MLC. Notably, the partial RDA  
409 shows that the underlying environmental changes (represented by depth and LOI) are more  
410 influential on the diatom assemblage than any tephra effects.

411 Nevertheless, there is evidence for a limited tephra effect and there are several hypotheses  
412 regarding the potential nature of tephra impacts on lake ecosystems, all of which have been  
413 reported in other tephra impact studies; 1) acidification in response to dry deposition of  
414 H<sub>2</sub>SO<sub>4</sub> following tephra deposition (Blinman *et al.*, 1979; Birks and Lotter, 1994; Bradbury *et*  
415 *al.*, 2004); 2) change in the nutrient status of the lake following tephra deposition (Barker *et*  
416 *al.*, 2003; Telford *et al.*, 2004); 3) habitat change following tephra deposition (Brant and  
417 Bahls, 1995; Telford *et al.*, 2004).

418

### 419 **Tephra and acidification**

420

421 It is important to explore the acidification hypothesis, as this is one of the most commonly  
422 reported impacts of tephra deposition (Blinman *et al.*, 1979; Birks and Lotter, 1994; Bradbury  
423 *et al.*, 2004). In MLF (and to a lesser extent MLC) increases of *Brachysira brebissonii*,  
424 *Tabellaria flocculosa*, *Frustulia rhomboides* and *Eunotia naegeli* potentially support the  
425 acidification hypothesis as these are acid indicators (Anderson and Renberg, 1992; Fránková  
426 *et al.*, 2009) and the declines of the more acid sensitive *Fragilaria brevistriata* and *Staurosira*

427 *venter* are also consistent with an acidification (Anderson and Renberg, 1992). Importantly,  
428 however, the diatom response in MLC across all taxa is not consistent with acidification. For  
429 example, *Craticula halophila* increases significantly after tephra deposition and prefers  
430 alkaline conditions (Round *et al.*, 1990).

431 Similarly the diatom data from MLC could be interpreted as showing support for the  
432 acidification hypothesis due to the significance of model 1, and associated changes in  
433 *Fragilaria brevistriata* and *Staurosira venter* consistent with increased acidity following  
434 tephra deposition. However, the associated increases in *Nitzschia palea*, *Aulacoseira* species  
435 and *Discostella pseudostelligera* are counter to expected floristic trends following  
436 acidification, as they should decrease with increased acid loading (Anderson and Renberg,  
437 1992; Saros and Anderson, 2015). We therefore find no clear evidence of post-tephra  
438 acidification in this oligotrophic lake.

439

#### 440 **Tephra and nutrient status**

441

442 The increase of *Aulacoseira* taxa in MLC is consistent with a change in nutrient status, as  
443 *Aulacoseira* taxa thrive under silica rich conditions (Thwaites, 1848; Abella, 1988; Caballero  
444 *et al.*, 2006). The increase of silica at that time is most likely to be from tephra deposition,  
445 which also prevents the regeneration of nutrients such as phosphorus as it creates an  
446 impermeable barrier over the lakes sediment (Barker *et al.*, 2000; 2003). The reduction of  
447 phosphorous may have been a limiting factor for *Fragilaria brevistriata* and *Staurosira*  
448 *venter* (Abella, 1988), but their generally high tolerance suggests habitat change could have  
449 also been influential (discussed below). This hypothesis of a nutrient change is also supported

450 by the partial RDA analyses (Table 3) for MLC model 1, as *Aulacoseira crassipunctata* and  
451 other *Aulacoseira* taxa respond positively to tephra deposition, and thus the influx of silica.  
452 Likewise, *Fragilaria brevistriata* and *Staurosira venter* respond negatively in response to  
453 tephra in the partial RDA analyses, which might also suggest a habitat change (Table 3).

454

#### 455 **Tephra and habitat change**

456

457 The floristic changes (increase of *Aulacoseira alpigena* and *Aulacoseira lirata*, decrease of  
458 *Fragilaria brevistriata* and *Staurosira venter* and increase of epipellic taxa) coincident with  
459 tephra deposition in MLC (and MLF despite the insignificance of tephra) may also suggest an  
460 alteration of habitat. The decrease of tychoplanktonic *Fragilaria brevistriata* and *Staurosira*  
461 *venter* and epiphytic taxa are indicative of disturbance and habitat change as aquatic plants  
462 are likely to be adversely affected by tephra deposition due to blanket burial. Although  
463 *Fragilaria brevistriata* and *Staurosira venter* are tychoplanktonic (Caballero *et al.*, 2006),  
464 they have been reported as epiphytic (Ehrlich, 1995; Stoermer *et al.*, 1996) due to their wide  
465 tolerances and tendency to attach themselves to benthic aquatic plants following a  
466 disturbance in the tychoplanktonic zone (e.g. tephra deposition). As such, these species often  
467 fluctuate along with aquatic vegetation (Caballero *et al.*, 2006). The local vegetation record  
468 from pollen analysis at Moss Lake (Egan *et al.*, 2016), summarised in Figure 7, shows a  
469 decrease of *Myriophyllum spicatum* and *Nuphar* due to blanket coverage inhibiting gas  
470 exchange and photosynthesis, which would result in a reduction in aquatic vegetation  
471 (submerged and emergent) thus epiphytic species and *Fragilaria brevistriata* and *Staurosira*  
472 *venter* would decline due to habitat loss (Telford *et al.*, 2004).

473 The increase of epipelagic taxa, *Staurosira lapidicola*, *Sellaphora pupula* and *Brachysira*  
474 *brebissonii* further demonstrates habitat change as these species increase in response to tephra  
475 deposition through colonisation of the newly deposited tephra (Telford *et al.*, 2004). The  
476 increase of *Aulacoseira* taxa also suggests habitat alteration as they can thrive in turbid  
477 waters with low light (Caballero *et al.*, 2006). Partial RDA confirmed tephra to be a  
478 significant variable in model 1, with *Staurosira venter* and *Fragilaria brevistriata* responding  
479 negatively to the tephra and the potential decline in suitable habitat as a result, and *Nitzschia*  
480 *palea* responding positively to the tephra due to the increase in habitat availability, further  
481 supporting this hypothesis.

482

### 483 **Impact summary**

484

485 In summary, tephra from Mount Mazama had a significant unique effect on the diatom  
486 assemblage of the central core (MLC) according to partial RDA model 1, but no significant  
487 effect in the fringe core (MLF), or the central core (MLC) when applying partial RDA  
488 models 2 and 3. There is therefore some evidence of a tephra effect but this is inconsistent.  
489 We argue the nature of the effect is a change of habitat conditions and an increase in the Si:P  
490 ratio. These impacts lasted for 150-200 years (based on zone data, PCA axis 2 and the age  
491 depth model) to a maximum of 500 years with most recovery having happened within 200-  
492 250 years (based on partial RDA model one). The transition from diatom zone C2 to C3  
493 reflects the point where the diatom assemblage reverts back to pre-tephra conditions, which is  
494 also evident from PCA axis 2 (Figure 5). This is demonstrated by increases of *Fragilaria*  
495 *brevistriata* and *Staurosira venter* to baseline levels, reflecting the 150-200 year recovery  
496 period. This recovery time period is nearly consistent with partial RDA model 1, which

497 implies a maximum recovery period of 500 years, but with most recovering happening within  
498 200-250 years. Importantly all partial RDA models indicate that depth and LOI had  
499 significant unique effects. This suggests other drivers of change in the aquatic system at the  
500 time of the Mazama event, reflecting wider changes in climate, sedimentology and  
501 vegetation.

502

### 503 **Alternative drivers of change**

504

505 All partial RDA analyses for MLC and MLF (Table 3) indicate that depth (surrogate for  
506 directional change) had the most significant unique effect on the diatom assemblage (MLF,  
507 37.7%; MLC, 10.6-12.7%). LOI was also identified as an important variable (MLF, 15.5%;  
508 MLC, 9.1-11.3%). Taxa explained by depth in MLF are *Staurosira venter*, *Nitzschia palea*,  
509 *Brachysira brebissonii*, *Tabellaria flocculosa*, and *Aulacoseira lirata*. These taxa represent  
510 the most notable changes in the MLF diatom assemblage, and although the partial RDA  
511 analysis indicates depth to have the greatest significant unique effect there are a few caveats  
512 with the partial RDA analysis that must be highlighted. There is a potential issue with the  
513 MLF model as there was no evidence of recovery, which is what the model assumed. Given  
514 the sensitivity of lake fringes one might have expected a more pronounced impact of tephra  
515 here. Instead depth appears to be the significant variable but there is a possibility that depth  
516 could have been acting as a surrogate for tephra (as there is no evidence of recovery) and in  
517 fact tephra may be more important than indicated by this model. Despite this, all but one  
518 model for MLC indicate an insignificant independent response to tephra deposition, with  
519 depth (directional change) being the main driver.

520

521 Taxa explained by depth in MLC are *Aulacoseira crassipunctata* and *Aulacoseira lirata* in all  
522 models and *Discostella pseudostelligera* in model one (Table 3). *Discostella pseudostelligera*  
523 are likely to be responding to ongoing climate change as PCA axis 1 (Figure 5) is  
524 representative of short term environmental change and correlates well with variations in  
525 *Discostella pseudostelligera*. Further, Egan (2016) report fluctuations of *Aulacoseira* taxa  
526 and *Discostella pseudostelligera* in response to climatic changes at that time. LOI also had a  
527 significant unique effect and is likely to reflect the change in sedimentology around the time  
528 of Mazama tephra deposition from silty gyttja to organic peaty silt (Figure 2). This change in  
529 sedimentology could potentially be as a result of tephra deposition as it can reduce infiltration  
530 and increase surface wetness potentially creating waterlogged conditions for peaty silt to  
531 develop, however, as models 2 and 3 for MLC and MLF indicate an insignificant effect of  
532 tephra, this is unlikely. Alternatively, LOI could be influenced by longer-term environmental  
533 change. During the time of tephra deposition Moss Lake was in a period of transition with an  
534 increase of nutrients from both a warmer climate (beginning 8000 cal yr BP until 6500 cal yr  
535 BP) and the developing conifer forest ecosystem, particularly *Pseudotsuga menziesii*, *Tsuga*  
536 *heterophylla* and Cupressaceae, at that time (Egan *et al.*, 2016), summarised in Figure 7b.  
537 There is also evidence that during the Holocene the lake was undergoing hydroseral  
538 succession in the marginal areas of the lake basin as evidenced by the increasing LOI and  
539 development of silty peats (Figure 6). These catchment wide changes allowed longer growing  
540 seasons, increased nutrient availability and increased habitat availability for diatoms,  
541 increasing diatom diversity (Egan, 2016). Thus the addition of tephra may have amplified  
542 these changes in nutrient status and hydroseral succession as the addition of silica into the  
543 system would have a positive effect on some diatoms (i.e. *Aulacoseira* taxa), while the tephra  
544 itself further increases habitat availability for the epipellic species and contributes to sediment

545 influx. However, given that the partial RDA analyses only indicate that tephra had a  
546 significant unique effect in one model it can be concluded that the effect of tephra on the  
547 aquatic system of Moss Lake was minimal in relation to underlying environmental trends.

548

## 549 **CONCLUSION**

550

551 The climactic eruption of Mount Mazama was a major volcanic event in North America  
552 during the Holocene. At Moss Lake up to 50 mm of tephra was deposited. There is some  
553 evidence for significant, and independent impacts of this tephra on the aquatic ecosystem, but  
554 these are not consistent between the central and fringe core locations. Partial RDA analyses  
555 indicated that tephra had no unique significant effect on the diatom assemblages of the fringe  
556 core location, whereas the partial RDA models for the central core location based on a 500  
557 year maximum recovery period, with most recovery happening with 200-250 years, were  
558 significant.

559 The diatom response recorded in both the shallow and deep water lake systems suggests that  
560 there was a change in habitat availability with a reduction in tychoplanktonic taxa  
561 (particularly *Fragilaria brevistriata* and *Staurosira venter*) and epiphytic taxa, and an  
562 increase in epipelagic taxa (particularly *Brachysira brebissonii* and *Frustulia rhomboides*).

563 There was also a change in the Si:P ratio with an increase of *Aulacoseira* taxa recorded in the  
564 deep water lake system. These changes are coincident with the timing of tephra deposition  
565 and are also associated with ongoing environmental change within the catchment, with both a  
566 warmer climate and the expansion of a conifer forest evidenced by pollen analyses on MLF  
567 and MLC. The climate and catchment wide changes could have been amplified by tephra



568 deposition due to the addition of silica contributing to nutrient availability and the sediment  
569 (tephra) influx increasing epipelagic habitat availability.

570 Overall the partial RDA analyses indicate some evidence of an effect, likely to be habitat  
571 change, but this is not consistent between the central and fringe core and the tephra impact is  
572 not as important as changes in the assemblages caused by underlying environmental trends.  
573 There is a natural tendency to equate any coincidental diatom change with the impact of  
574 tephra deposition. Without high resolution analyses, cross correlations with multiple cores  
575 and other records, and robust statistical analyses, it is difficult to determine how influential  
576 tephra is.

577

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585

## 586 **SUPPLEMENTARY AND ARCHIVED DATA**

587

588 All diatom and stratigraphic data are available from:  
589 <https://doi.pangaea.de/10.1594/PANGAEA.890666> in addition to the supplementary data  
590 already discussed within the manuscript.

591

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750

751 Table 1: Conventional (<sup>14</sup>C yr BP), calibrated (cal yr BP) and modelled (at 95.4% probability  
752 range) radiocarbon ages for MLC previously reported in Egan *et al.*, (2016).

Lab no.	Depth (cm)	Material	Age ( <sup>14</sup> C years BP ± 1 SD)	Age range (cal yr BP 2 SD)	Modelled age range (cal yr BP 95.4% probability range)
<b>SUERC-59476</b>	305	Organic sediment	6330 ± 36	7410-7167	7286-7163
<b>SUERC-59477</b>	315	Organic sediment	6590 ± 38	7565-7430	7486-7424
<b>SUERC-59478</b>	321	Organic sediment directly above Mazama	6687 ± 39	7619-7480	7622-7531
<b>Mazama*</b>	324	-	-	7682-7584*	
<b>SUERC-59479</b>	345	Organic sediment	7430 ± 39	8344-8180	8351-8178

753 \*Age range from Egan et al. (2015)

754 \*\* Age range based on deposition model

755 Table 2: Summary of the diatom assemblage from MLC and MLF, and their associated  
 756 zones.

Zone	Depth (cm)	Diatom description	Diatom concentration
<b>MLC</b>			
C3	316.5	<ul style="list-style-type: none"> <li>- <i>Discotella pseudostelligera</i> decrease to 50% but remain dominant.</li> <li>- Tycho planktonic species dominate the benthic community (up to 50%).</li> </ul>	Variable ( $2 \times 10^8$ - $10 \times 10^7$ per g dry weight).
C2	325	<ul style="list-style-type: none"> <li>- Planktonic and tycho planktonic species continue to dominate (90%).</li> <li>- <i>Aulacoseira crassipunctata</i> appear in greater abundance (10%).</li> <li>- <i>Fragilaria brevistriata</i> and <i>Staurosira venter</i> start to increase from 5% to 20%.</li> <li>- Epipelagic species <i>Nitzschia palea</i>, <i>Navicula bremensis</i> and <i>Sellaphora pupula</i> briefly increase.</li> </ul>	Variable ( $1 \times 10^8$ - $13 \times 10^7$ per g dry weight).
C1	340	<ul style="list-style-type: none"> <li>- Planktonic <i>Discotella pseudostelligera</i> dominate (up to 60%).</li> <li>- Upon tephra deposition tycho planktonic <i>Aulacoseira</i> species increase but <i>Fragilaria brevistriata</i> and <i>Staurosira venter</i> decrease.</li> </ul>	Increases upon tephra deposition.
<b>MLF</b>			
F2	156.8	<ul style="list-style-type: none"> <li>- Epipelagic species become increasingly important, especially <i>Brachysira brebissonii</i>, <i>Craticula halophila</i> and <i>Nitzschia palea</i>, which increase during tephra deposition</li> <li>- Epiphytic species modestly decline following tephra deposition.</li> <li>- Tycho planktonic species decrease after tephra deposition, <i>Staurosira venter</i> nearly disappears.</li> </ul>	Low with a maximum of $1 \times 10^8$ per g dry weight.

F1	160	<ul style="list-style-type: none"> <li>- Tychoplanktonic species fluctuate from 10% to 50% decreasing just before tephra deposition.</li> <li>- Epiphytic species dominate just before and upon tephra deposition (~40%).</li> <li>- <i>Discotella pseudostelligera</i> briefly appears before tephra deposition and declines towards the top of the zone.</li> <li>- <i>Tabellaria flocculosa</i> increases just before tephra deposition, then declines.</li> </ul>	Variable ( $0.5 \times 10^8$ - $7 \times 10^8$ per g dry weight).
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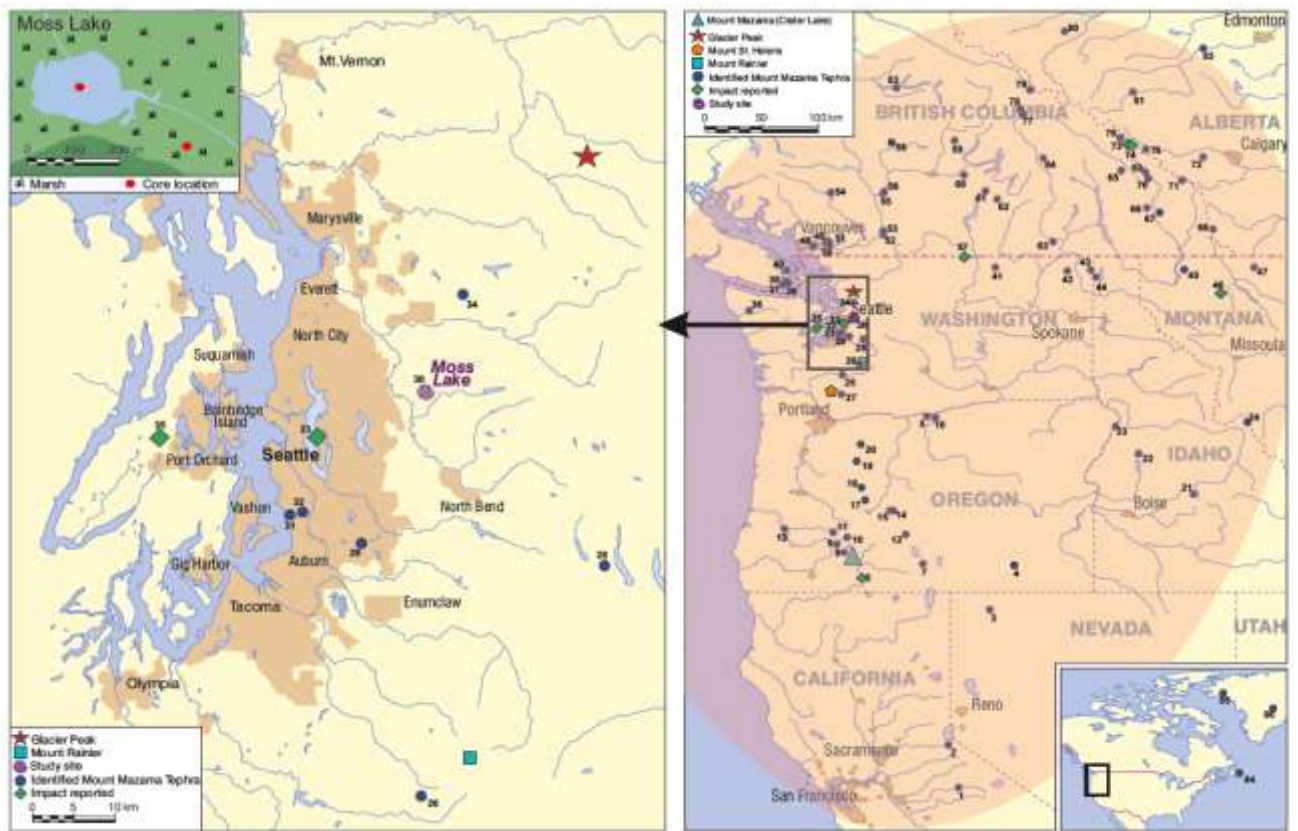
761

762 Table 3: Results of partial redundancy analysis of the diatom stratigraphical data sets of Moss  
 763 Lake central and Moss Lake fringe reporting the unique effects and their significance. Those  
 764 in **bold** are significant. Diatom species with an abundance of >5% or present in at least 10  
 765 samples was used in the analysis. Lower down on the table is the percentage variation of the  
 766 diatoms, which indicates which species are most influenced by the variables. The +/- signs  
 767 means the species had either a positive or negative relationship with that particular variable.

Variable	Tephra	Depth	LOI
Co-variables	Depth + LOI	Tephra + LOI	Tephra + Depth
<b>Moss Lake Fringe (MLF)</b>			
Unique effect (%)	3.6	<b>37.7</b>	<b>15.4</b>
Significance of unique effect	0.204	<b>0.01</b>	<b>0.031</b>
<b>Moss Lake Central (MLC)- Model 1</b>			
Unique effect (%)	<b>11.2</b>	<b>10.6</b>	6.1
Significance of unique effect	<b>0.02</b>	<b>0.02</b>	0.103
<b>Moss Lake Central (MLC)- Model 2</b>			
Unique effect (%)	7.1	<b>12.7</b>	<b>9.1</b>
Significance of unique effect	0.059	<b>0.037</b>	<b>0.048</b>
<b>Moss Lake Central (MLC)- Model 3</b>			
Unique effect (%)	4.5	<b>12.6</b>	<b>11.3</b>
Significance of unique effect	0.252	<b>0.046</b>	<b>0.031</b>
<b>% Variation of Response Variable (Diatoms)</b>			
<b>Tephra</b>	<b>Depth</b>	<b>LOI</b>	
Moss Lake Fringe			
<i>Eunotia obliquistriata</i> (+7.8), <i>Eunotia macroglossa</i> (-7.8), <i>Stauroneis lapidicola</i> (-7.5), <i>Eunotia arcus</i> (+6.6), <i>Navicula bremensis</i> (-5.1)	<i>Staurosira venter</i> (+56.2), <i>Nitzschia palea</i> (-55.2), <i>Brachysira brebissonii</i> (-50.2), <i>Tabellaria flocculosa</i> (-47.0), <i>Aulacoseira lirata</i> (+42.6)	<i>Aulacoseira alpigena</i> (+23.1), <i>Brachysira brebissonii</i> (-18.5), <i>Gomphonema gracile</i> (+17.5), <i>Frustulia rhomboides</i> (-17.0), <i>Stauroneis kreigeri</i> (+13.6)	
Moss Lake Central- Model 1			
<i>Staurosira venter</i> (-65.3), <i>Fragilaria brevistriata</i> (-47.5), <i>Aulacoseira crassipunctata</i> (+44.9), <i>Nitzschia palea</i> (+36.0), <i>Discostella pseudostelligera</i> (+31.1)	<i>Aulacoseira crassipunctata</i> (-29.8), <i>Aulacoseira lirata</i> (-15.3), <i>Discostella pseudostelligera</i> (+15.2), <i>Aulacoseira alpigena</i> (-7.8)	<i>Aulacoseira crassipunctata</i> (-13.2), <i>Aulacoseira granulata</i> (-12.2), <i>Staurosira venter</i> (+6.5)	
Moss Lake Central- Model 2			
<i>Staurosira venter</i> (-55.8), <i>Aulacoseira crassipunctata</i> (+42.5), <i>Fragilaria brevistriata</i> (-41.6), <i>Aulacoseira valida</i> (+24.5), <i>Nitzschia palea</i> (+23.8)	<i>Aulacoseira crassipunctata</i> (-38.4), <i>Aulacoseira lirata</i> (-19.2), <i>Pseudostaurosira elliptica</i> (+10.0), <i>Staurosira venter</i> (+9.6)	<i>Aulacoseira crassipunctata</i> (-19.8), <i>Staurosira venter</i> (+16.7), <i>Fragilaria brevistriata</i> (+11.8), <i>Aulacoseira granulata</i> (-9.4)	
Moss Lake Central- Model 3			
<i>Aulacoseira valida</i> (+19.0), <i>Staurosira venter</i> (-17.4), <i>Fragilaria brevistriata</i> (-9.0)	<i>Aulacoseira crassipunctata</i> (-38.0), <i>Aulacoseira lirata</i> (-18.5), <i>Staurosira venter</i> (+12.0), <i>Pseudostaurosira elliptica</i> (+10.6)	<i>Staurosira venter</i> (+26.9), <i>Aulacoseira crassipunctata</i> (-21.7), <i>Fragilaria brevistriata</i> (+19.1), <i>Aulacoseira pusilla</i> (-10.4), <i>Aulacoseira granulata</i> (-10.1)	

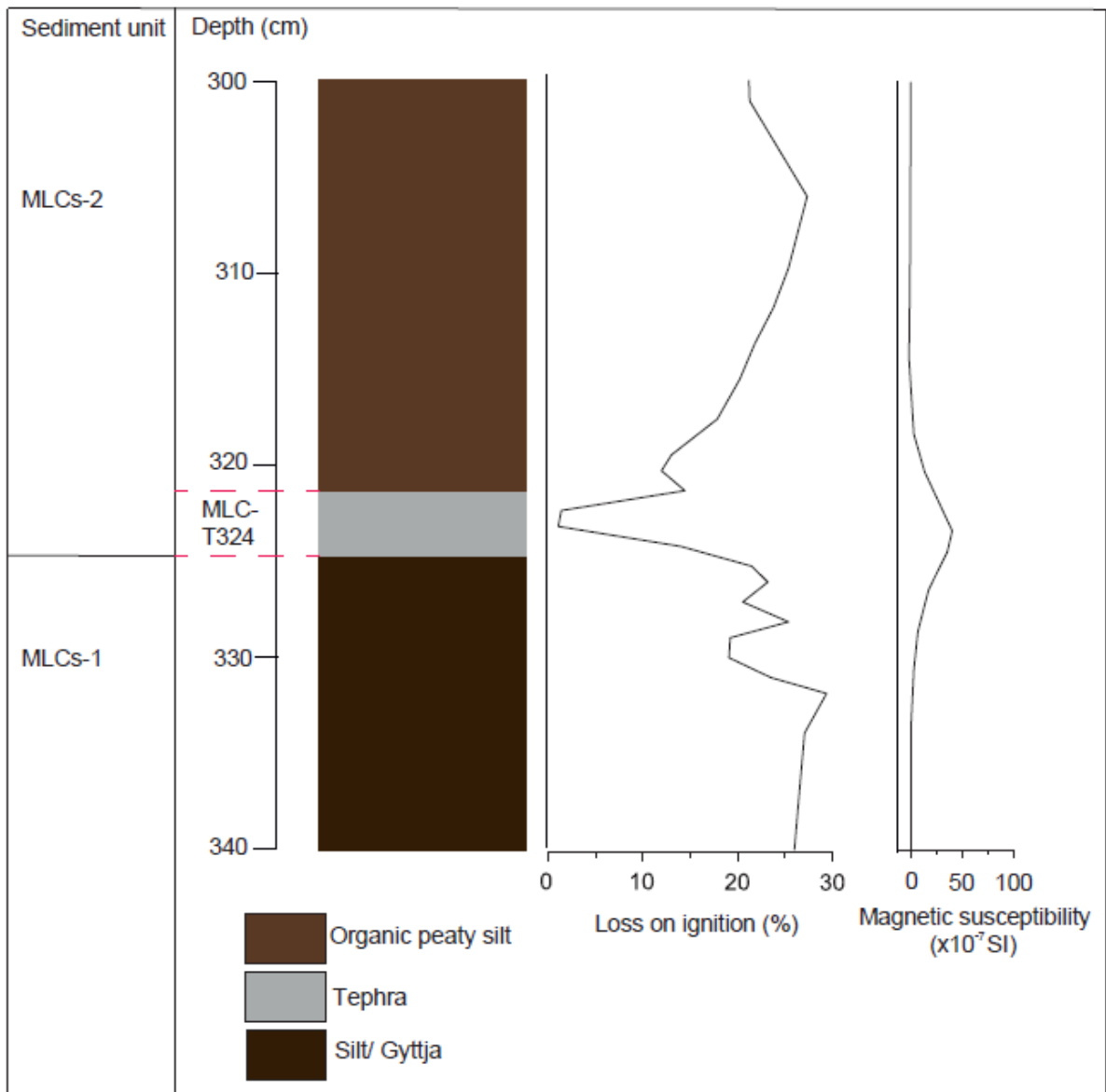
770 **FIGURES**

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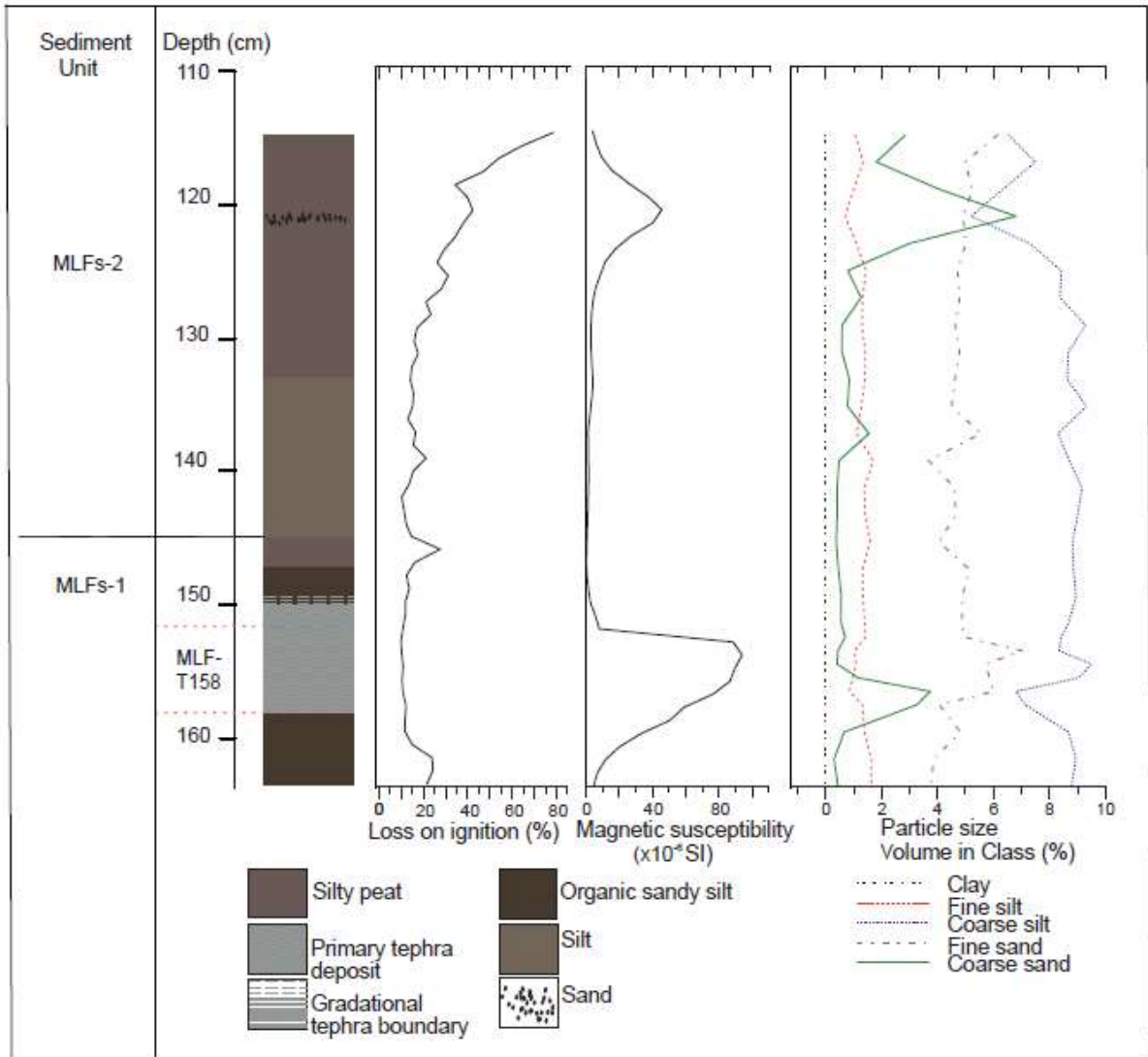
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773 Figure 1: Extent of deposition from the Plinian eruption of Mount Mazama, and sites where it  
774 has previously been identified. The elliptical shaded envelope in the map to the right shows  
775 the extent of recorded visible Mount Mazama tephra deposition. True tephra dispersal was  
776 much greater with cryptotephra having been found as far as Newfoundland (Pyne-O'Donnell  
777 et al., 2012) and Greenland (Zdanowicz et al., 1999). The locations of Moss Lake, Mount  
778 Mazama and Glacier Peak are also highlighted. The shading around cities indicates the size  
779 and distribution of major urban areas. A key is provided for the numbered sites in  
780 supplementary material Table 2.



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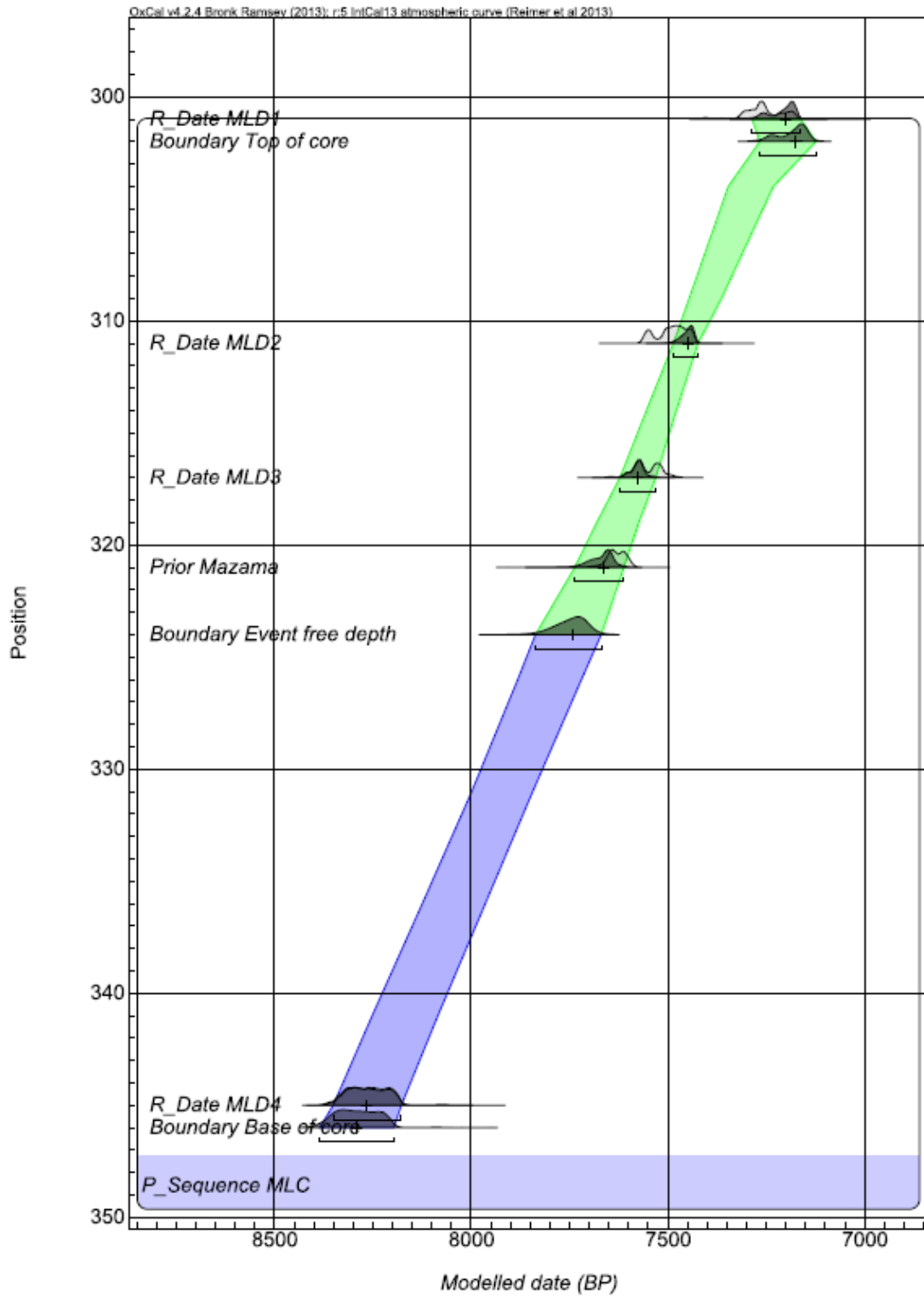
782 Figure 2: Lithology, % LOI, magnetic susceptibility and carbonate content of MLC.



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784 Figure 3: Lithology, % LOI, magnetic susceptibility, carbonate content and particle size of

785 MLF



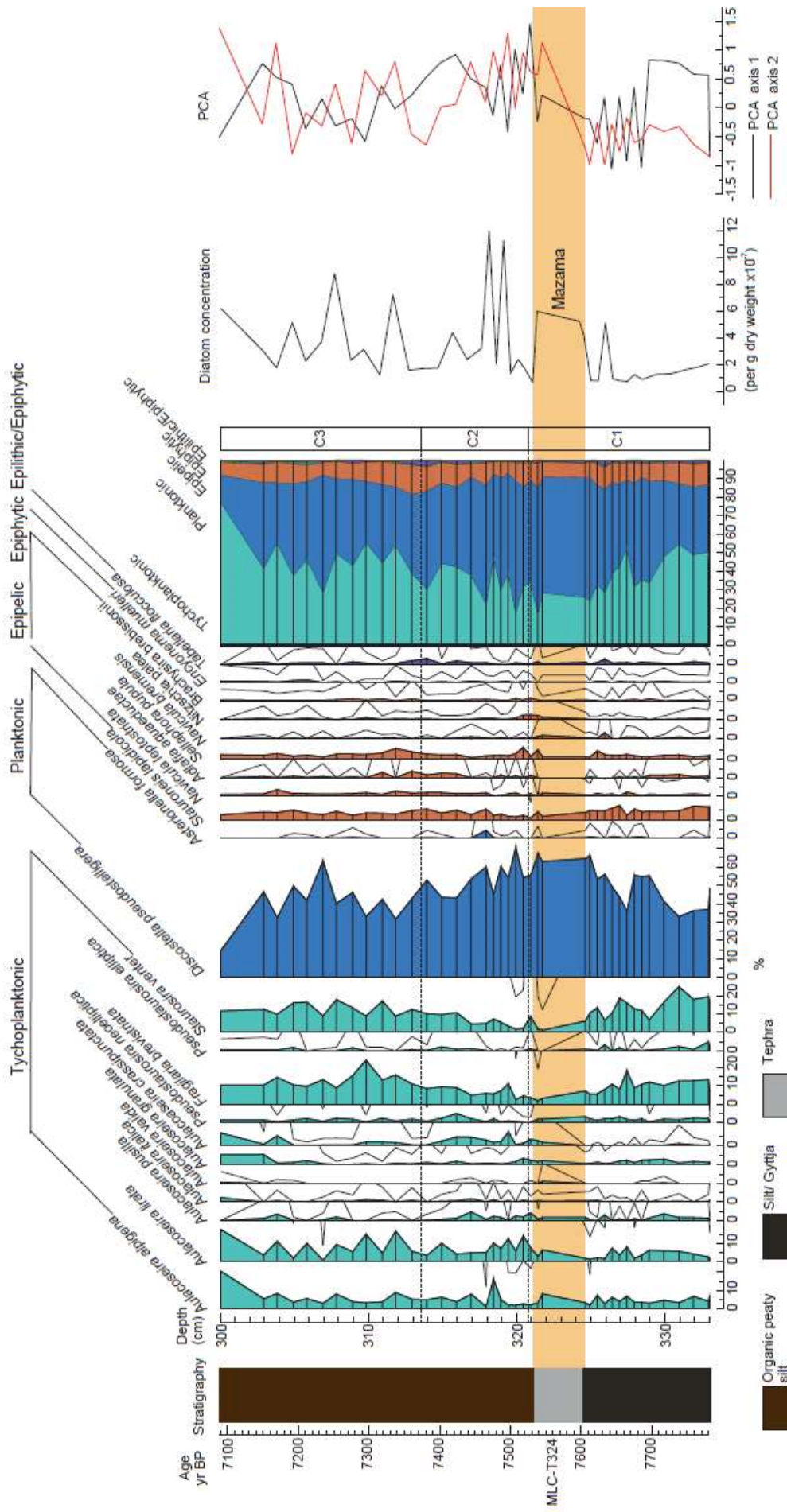
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787 Figure 4: Bayesian age-depth (OxCal v.4.2 (Bronk Ramsey, 2014)) model for MLC derived

788 from the comparison of the radiocarbon ages calibrated using the IntCal13 (Reimer, 2013)

789 dataset.





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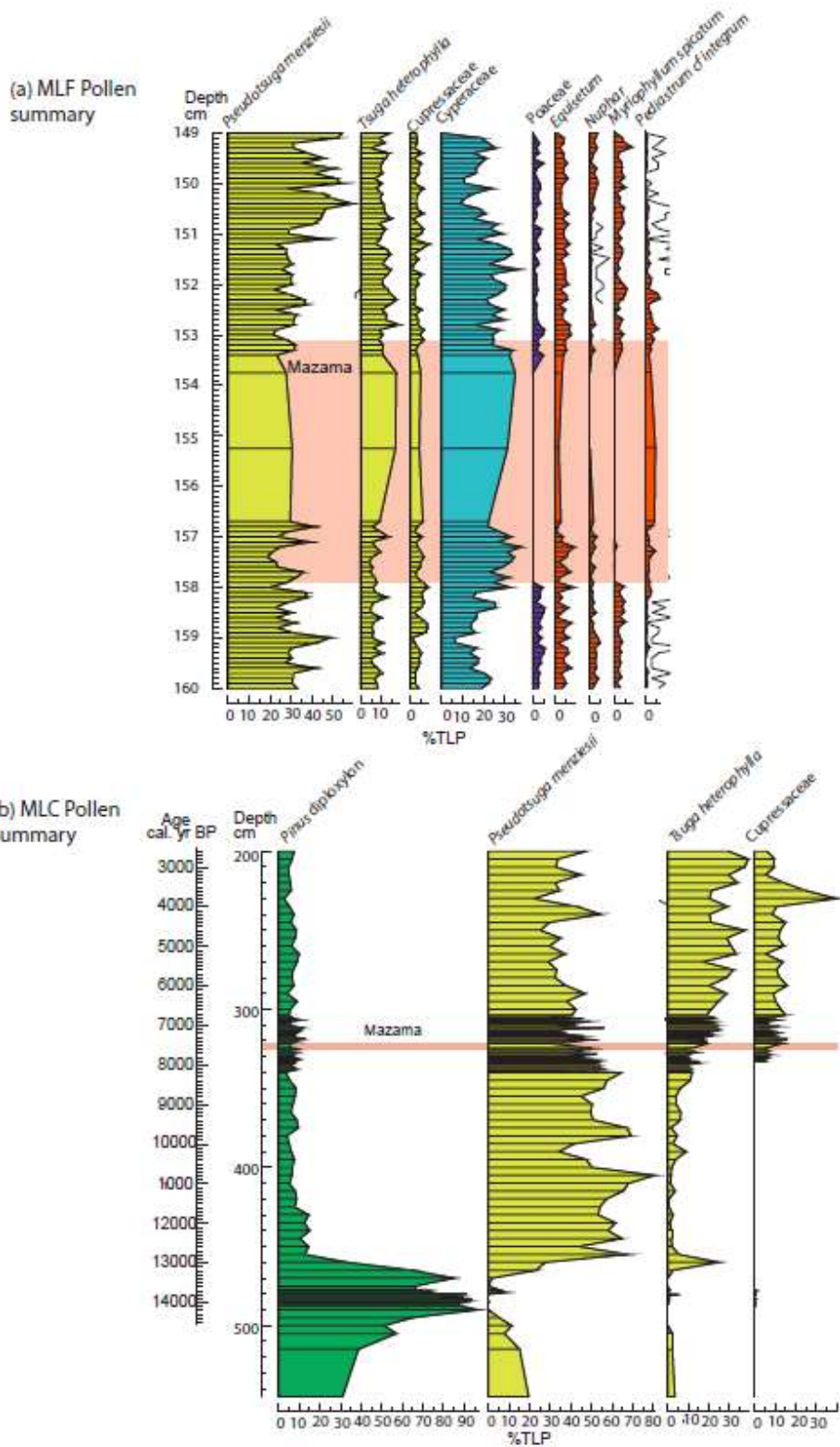
791 Figure 5: Diatom assemblage from Moss Lake central displaying the lithology, percentage of  
792 diatoms, summary diagram, diatom zonation, diatom concentration and PCA axis 1 and 2.  
793 The shaded bar represents the location of the Mazama tephra (MLC-T324), also labelled. The  
794 solid line on percentage diagram is 10x exaggeration.

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796 Figure 6: Diatom assemblage from Moss Lake fringe displaying the lithology, percentage of  
797 diatoms, summary diagram, diatom zonation, diatom concentration and PCA axis 1 and 2.  
798 The shaded bar represents the location of the Mazama tephra (MLF-T158), also labelled. The  
799 solid line on percentage diagram is 10x exaggeration.

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801 Figure 7: Summary pollen diagrams from (a) Moss Lake Fringe (MLF) and (b) Moss Lake  
 802 Central (MLC). Full pollen diagrams are presented in Egan *et al.*, (2016). Species coloured:  
 803 green= xerophytes, yellow= mesophyte, blue= hydrophyte, orange= spores and aquatic.

## Supplementary Material

Table 1: Geographical coordinates, topographical and limnological properties, water temperature, pH, conductivity and water chemistry for Swamp Lake and Moss Lake. Water chemistry includes the concentrations of: Total Organic Carbon (TOC), Total Carbon (TC), Inorganic Carbon (IC), Total Nitrates (TN), Chlorides, Nitrites, Nitrates, phosphates, Sulphates, Aluminium (Al), Calcium (Ca), Copper (Cu), Iron (Fe), Potassium (K), Lithium (Li), Magnesium (Mg), Manganese (Mn), Sodium (Na), Nickel (Ni), Lead (Pb) and Zinc (Zn).

Variable	Units	Moss Lake	
Latitude	(N)	47° 41' 35.7"	
Longitude	(w)	121° 50' 48.6"	
Distance from Mazama	(km)	530	
Altitude	(m asl)	158	
Max depth	(m)	4.5	
Area (approx.)	(m <sup>2</sup> )	13,275	
		July 2013	May 2014
pH		6.15	6.3
Conductivity	( $\mu\text{S cm}^{-1}$ )	14	22
Water temp	(°C)	-	18.3
TOC	(ppm)	15.07	-
TC	(ppm)	17.12	-
IC	(ppm)	2.049	-
TN	(ppm)	0.4697	-
Chloride	(ppm)	2.284	3.834
Nitrite	(ppm)	-	0.029
Nitrate	(ppm)	-	0.353
Phosphate	(ppm)	0.849	-
Sulphate	(ppm)	0.146	0.458
Al	(ppm)	0.28	0.163
Ca	(ppm)	4.035	2.741
Cu	(ppm)	0.011	0.012
Fe	(ppm)	0.281	0.061
K	(ppm)	1.692	0.046
Li	(ppm)	0.041	0.023
Mg	(ppm)	0.969	0.694
Mn	(ppm)	0.002	0.003
Na	(ppm)	2.254	1.412
Ni	(ppm)	0.002	0.005
Pb	(ppm)	0.011	0.01
Zn	(ppm)	0.253	0.02

Table 2. Locations and references of point provided in Figure 1 of the main document.

<b>Point</b>	<b>Site name</b>	<b>Reference</b>
1	Swamp Lake	(Street et al., 2012)
2	Osgood Swamp	(Adam, 1967)
3	Virgin Creek	(Davis, 1978)
4	Wildhorse Lake	(Blinman et al., 1979)
5	Wildcat Canyon	(Randle et al., 1971)
6	Upper Klamath Lake	(Bradbury et al., 2004)
7	Paisley Cave	(Preston et al., 1955)
8	Crater Lake Vicinity	(Bacon, 1983)
9	Sparks Lake	(Kittleman, 1973)
10	Diamond Lake	(Kittleman, 1973)
11	Toketee Falls	(Rubin & Alexander, 1960)
12	Fort Rock Cave	(Randle et al., 1971)
13	North Umpqua River Valley	(Bacon, 1983)
14	Paulina Lake	(Kittleman, 1973)
15	East Lake	(Kittleman, 1973)
16	Hobo Cave	(Randle et al., 1971)
17	Tumalo Lake	(Long et al., 2014)
18	Three Creek	(Long et al., 2014)
19	Round Lake	(Long et al., 2014)
20	Breitenbush Lake	(Long et al., 2014)
21	Lower Decker Lake	(Whitlock et al., 2011)
22	McCall Fen	(Doerner & Carrara , 2001)

- 23 Muir Creek (Arnold and Libby, 1951; Crane, 1956; Kittleman, 1973; Valastro et al., 1968)
- 24 Lost Trail Pass Bog (Blinman et al., 1979; Mehringer et al., 1977a)
- 25 Mount Rainer National Park (Mullineaux, 1974)
- 26 Bear Swamp (Blackford, Pers. Comm)
- 27 Davis Lake (Barnosky, 1981)
- 28 Swamp Lake (Blackford pers comm; Egan, Unpublished)
- 29 Covington (Broecker et al., 1956)
- 30 Moss Lake This paper
- 31 Bow Lake (Rubin & Alexander, 1960)
- 32 Arrow Lake (Rubin & Alexander, 1960)
- 33 Lake Washington (Abella, 1988; Leopold et al., 1982)
- 34 Skykomish River (Tabor et al., 1963)
- 35 Wildcat Lake (Blinman et al., 1979)
- 36 Bogachiel River (Heusser, 1983)
- 37 Rithets Bog (Lowdon & Blake, 1970)
- 38 Pike Lake (James et al., 2009)
- 39 Maltby Lake (James et al., 2009)
- 40 Portage Inlet (Buckley & Willis, 1970)
- 41 Bonaparte Meadows (Mack et al., 1979)
- 42 Big Meadow Lake (Powers & Wilcox, 1964)
- 43 Huff Lake (Moseley et al., 1992)
- 44 Hager Lake (Moseley et al., 1992)
- 45 Tepee Lake (Mack et al., 1983)
- 46 Foy Lake (Power et al., 2011)



47	Swiftcurrent Lake	(MacGregor et al., 2011)
48	Burnaby Lake	(Dyck et al., 1966)
49	Lake Mike	(Brown et al., 1989)
50	Marion Lake	(Mathewes, 1973)
51	Surprise Lake	(Mathewes, 1973)
52	Fraser Canyon	(Lowdon et al., 1969)
53	Squeah Lake	(Mathewes et al., 1972)
54	Lower Jaffre Lake	(Filippelli et al., 2006)
55	Drynoch Slide	(Sanger, 1967)
56	Drynoch Slide	(Lowdon et al., 1969)
57	Kilpoola Lake	(Heinrichs et al., 1999)
58	Green Lake	(Filippelli et al., 2006)
59	Dunn Peak	(Duford & Osborn, 1978)
60	Chase	(Lowdon & Blake, 1973)
61	Lavington	(Lowdon & Blake, 1970)
62	Deep Creek	(Dyck et al., 1965)
63	Lower Arrow Lake	(Dyck et al., 1965)
64	Mount Revelstoke	(Lowdon et al., 1971)
65	Cartwright Lake	(Beierle & Smith, 1998)
66	Copper Lake	(Beierle & Smith, 1998)
67	Johnson Lake	(Beierle & Smith, 1998)
68	Crowsnest Pass	(Driver, 1982)
69	Dog Lake	(Hallett et al., 1997)
70	Cobb Lake	(Hallett et al., 1997)
71	Upper Kananaskis Lake	(Beierle & Smith, 1998)

- 72 Frederick Lake (Beierle & Smith, 1998)
- 73 Mary lake (Hickman & Reasoner, 1994)
- 74 Opabin Lake (Hickman & Reasoner, 1994)
- 75 Copper Lake (White & Osborn, 1992)
- 76 Lake O'Hara (Hickman & Reasoner, 1994)
- 77 Columbia River Valley (Fulton, 1971)
- 78 Columbia River (Buckley & Willis, 1969)
- 79 Tonquinn Pass (Luckman et al., 1986)
- 80 Upper Pinto Fen (Yu, 2007)
- 81 Goldeneye Lake Fen (Yu, 2007)
- 82 Keephills Fen (Chagué-Goff et al., 1996)
- 83 Quesnel Lake (Gilbert & Desloges, 2012)
- 84 Nordans Pond Bog (Pyne-O'Donnell et al., 2012)
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- 86 GISP 2 (Zdanowicz et al., 1999)

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943 Table 3: Conventional ( $^{14}\text{C}$  years BP) and calibrated (cal. years BP) radiocarbon ages for

944 MLF.

Lab no.	Depth (cm)	Material	Age ( $^{14}\text{C}$ years BP $\pm$ 1 SD)	Age range (cal. years BP 2 SD)
SUERC-52705	147	Organic sediment	5645 $\pm$ 36	6496-6319
SUERC- 55693	147	Organic sediment	5796 $\pm$ 38	6713-6491
	(re-submission)			
SUERC-52704	151	Organic sediment directly above MLF-T158	4948 $\pm$ 37	5745-5599
SUERC-55690	151	Organic sediment directly above MLF-T158	5705 $\pm$ 35	6626-6407
	(re-submission)			
SUERC-52703	161	Organic sediment below MLF-T158	7049 $\pm$ 41	7958-7795

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