1 **Microstructure and carbon storage capacity of hydrated magnesium carbonates**

2 **synthesized from different sources and conditions**

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10

11 **Abstract**

12 Recently, the mineral carbonation via the reaction of CO_2 with saline aquafers received much

13 attention as one of the most promising ways for geologic CO_2 storage. This paper reports

14 microstructure and carbon storage capacity of hydrated magnesium carbonates (HMCs)

15 synthesized from different sources, *i.e.,* reject brine and commercial $Mg(OH)_2$ slurry, and under

16 different conditions, *i.e.,* pH (8-14) and $Mg(OH)_2$:CO_2 molar ratio (1:1-1:7). Results show that

17 dypingite $(Mg_5(CO_3)_4(OH)_2 \cdot 5H_2O)$ is the main phase forming at lower $Mg(OH)_2$:CO_2 ratios.

18 An increase in the $Mg(OH)_2$:CO_2 ratio and/or pH leads to the precipitation of nesquehonite

19 $(MgCO_3 \cdot 3H_2O)$. A unique "house of cards" texture, involving formation of the rosette-like

20 dypingite flakes on the surface of nesquehonite needles, is discovered under elevated pH and

21 $Mg(OH)_2$:CO_2 ratios. HMCs synthesized from reject brine exhibit a much higher carbon

22 storage capacity of 82.6% than that produced from the commercial $Mg(OH)_2$ slurry (43.7%).

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Findings from this study advance understanding of mineral recovery from reject brine and the capture and long-term storage of CO_2 in the form of HMCs.

1. Introduction

Fossil fuels have been the world's primary energy source, providing over 85% of the energy demands worldwide [1-6]. However, nearly 83% of the anthropogenic greenhouse gas (GHG) emissions are coming from combustion and non-fuel uses of fossil fuels [7]. CO_2, which is the main GHG, has caused most of the global warming since it has the highest positive radiative forcing and far more abundant in the atmosphere than other heat-trapping gases [8, 9]. The concentration of CO_2 in the atmosphere has increased ~30% from 325 parts per million (ppm) at the beginning of the industrial era in 1970 to 409.7 ppm in May 2017 measured in Mauna Loa Observatory. Consequently, much attention has been drawn to the carbon management [10-12]. Carbon capture and storage (CCS) provides a feasible way to reduce the build-up of CO_2 in the atmosphere [13, 14]. CCS concept covers broad fields such as ocean, terrestrial, geological, biological and chemical approaches to store CO_2 gas in the long term [15-18], among which mineral carbonation via the reaction of CO_2 with saline aquafers is one of the most promising geologic CO_2 storage options [19-23]. Magnesium-based minerals have attracted great attentions worldwide as they show the potential to sequestrate anthropogenic CO_2 to counterpart the global warming [24-27]. Furthermore, reactive magnesia (MgO) cement has been studied as a potential alternative to the Portland cement due to its ability to sequester significant amount of carbon dioxide (CO_2) and potential to be fully recycled [28-32].

48　Although the most thermodynamically stable carbonate for magnesium is in the anhydrous

49　form, *i.e.,* magnesite ($MgCO_3$) [33-35], formation of magnesite at the ambient condition is not

50　common. Instead, formation of hydrated magnesium carbonates (HMCs) prevails as Mg^{2+} ions

51　in the solution are highly hydrated [24]. HMCs are a class of magnesium compounds that form

52　in MgO-CO_2-H_2O systems, where the carbonation of magnesium systems generate a variety of

53　phases, including dypingite ($Mg_5(CO_3)_4(OH)_2 \cdot 5H_2O$) [36], hydromagnesite

54　($Mg_5(CO_3)_4(OH)_2 \cdot 4H_2O$) [37], and nesquehonite ($MgCO_3 \cdot 3H_2O$) [18, 38]. It has been reported

55　that inclusion of HMCs in reactive MgO cement enhances carbonation of the resulting binder
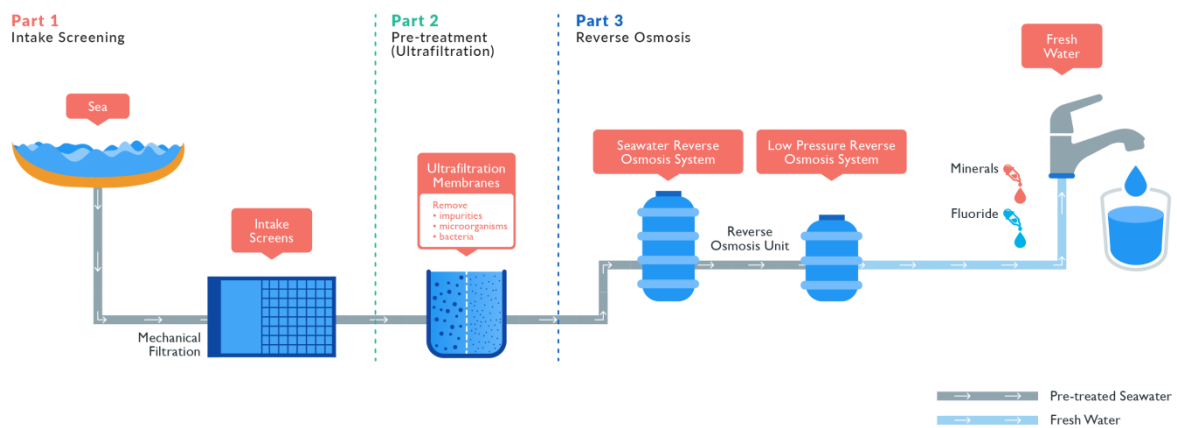
56　[39].

57

58　Theoretically, the formation of different phases of HMCs with different morphologies is

59　influenced by temperature, pH and CO_2 partial pressure [24, 40, 41]. Recent experimental

60　studies show that the precipitation of nesquehonite with needle-like morphology in an aqueous

61　solution occurs commonly under ambient conditions [15, 18, 38]. As the reaction temperatures

62　(333-368 K) and pH values increased, needle-like nesquehonite was replaced by

63　hydromagnesite with sheet-like morphology since nesquehonite is widely known to transform

64　to hydromagnesite at temperature above 50°C [40]. Hydromagnesite was reported to form at

65　120°C and P_{CO2} of 3 bar, which gradually transformed to magnesite within 5-15 hours.

66　However, a further increase of P_{CO2} to 100 bar at 120°C resulted in the direct precipitation of

67　magnesite [24]. The thermal behavior of the synthesized HMCs (*e.g.,* nesquehonite) has been

68　tested through real time in-situ X-ray diffraction (XRD), which indicated that nesquehonite and

69　dypingite remained thermal and structural stable up to 373 K and 435 K, respectively [36, 42].

70　Under the continued thermal treatment, nesquehonite transited into magnesite which was even

71　thermally stable up to 600 K, while dypingite transited into hydromagnesite at around 570 K,

72　assuring the long-term storage of CO_2. However, the influences of $Mg(OH)_2$/CO_2 molar ratios

73 and pH on the phases, morphology, and CCS efficiency of HMCs have yet been studied

74 systematically.

75

76 Desalination is a process that removes minerals from saline water. In coastal regions such as

77 Singapore where sources of fresh water are limited, desalination provides a feasible option to

78 produce fresh water. However, a high salty waste stream (*i.e.,* reject brine) would be generated

79 as much as the produced desalinated water at the end of the process [43]. Reject brine is a much

80 more complex media because chemicals are often added into the intake seawater (*e.g.,* to

81 precipitate the colloidal particles before running through the ultra-filtration) in the desalination

82 process (Fig. 1). Although many studies have investigated mineral trapping of CO_2 into saline

83 aquafers (*e.g.,* seawater, natural brine, or synthesized $MgCl_2$ solution) [18, 36, 42, 44-46], very

84 few has reported the use of reject brine as the CO_2 reservoir [47] and no study has proposed to

85 use reject brine as the $Mg(OH)_2$ source to synthesize HMCs.

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87

88 Figure 1 Schematic illustration of the typical process in a reverse osmosis desalination plant

89 [48]

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In this paper, influence of key parameters including $Mg(OH)_2$:CO_2 molar ratio, pH, and $Mg(OH)_2$ source on HMCs synthesis were investigated. The resulting HMCs were characterized by means of XRD, scanning electron microscopy (SEM), and thermogravimetric/differential thermal analysis (TG/DTA) to reveal phases, morphology, and CCS efficiency of HMCs synthesized under different conditions. In the following sections, materials and methodologies are presented first, followed by results presentation and discussion.

2. Materials and Methodology

2.1. Materials

Reject brine sample was collected from a local desalination plant which generates 318,500 m^3 desalinated water per day. The sample was filtrated through a 45 μm membrane filter to remove suspended solids before further analysis. An inductively coupled plasma-optical emission spectroscopy (ICP-OES) was used to analyze the chemical composition of reject brine in the current study (Table 1). Analytical grade $Mg(OH)_2$ (92% pure) and analytical grade sodium hydroxide (NaOH) with a purity of 97% were both purchased from VWR Pte Ltd in Singapore. Compressed CO_2 was purchased from Leeden National Oxygen Ltd in Singapore.

Table 1 Chemical composition of reject brine

Element	Cl	Na	SO$_4$	Mg	K	Ca	Sr	B	Si	Li	P	Al
Concentration (ppm)	65593	13580	4323	1718	846	471	14.6	3.8	3.7	0.3	0.2	0.1

2.2. Methodology

In the first approach, 0.82 g commercially available analytical grade $Mg(OH)_2$ was dissolved into 200 ml ultra-pure water to prepare the $Mg(OH)_2$ slurry. To study the influences of

114   $Mg(OH)_2$/CO_2 molar ratios and pH on the microstructures of HMCs , three sets of experiments

115   were designed at controlled conditions, *i.e.,* $Mg(OH)_2$:CO_2 molar ratio (1:1-1:7) and pH (8-14).

116   CO_2 was sparged into the slurry at a rate of 100 ml/min at room temperature under pre-

117   determined conditions as follows,

118   a) pH = 8, $Mg(OH)_2$:CO_2 molar ratio = 1:1 to 1:7

119   b) pH = 8 to 11, $Mg(OH)_2$:CO_2 molar ratio = 1:1

120   c) pH = 8 to 11, $Mg(OH)_2$:CO_2 molar ratio = 1:2

121

122   In the second approach, $Mg(OH)_2$ was first synthesized from the reject brine via the addition

123   of NaOH at an optimized condition (*i.e.,* NaOH/Mg^{2+} ratio of 2 at 25°C) determined from our

124   previous work [49-51], which results high yield and high purity $Mg(OH)_2$. After which, 0.82 g

125   of synthesized $Mg(OH)_2$ was mixed with 200 ml ultra-pure water. CO_2 was sparged into the

126   slurry at a rate of 100 ml/min at room temperature under controlled condition (*i.e.,* pH = 8,

127   $Mg(OH)_2$:CO_2 molar ratio = 1:1).

128

129   A pH/thermometer probe, inserted into reject brine, was used to continuously record the

130   temperature and pH during the experiment. A CO_2 flowmeter was used to monitor and record

131   the volume of CO_2 diffused into the slurry. Once the diffused CO_2 reached the pre-determined

132   value (*i.e.,* calculated based on the designed $Mg(OH)_2$:CO_2 molar ratio), the reaction was

133   terminated. 1M NaOH was added into the slurry continuously to maintain the pH to the

134   designed value since sparging CO_2 lowered the pH of slurry. HMCs were separated from the

135   solution by a centrifuge and washed three times using ultra-pure water. The washed samples

136   were fully dried in an oven at low temperature (*i.e.,* 40°C) to avoid any phase changes, before

137   being ground into powder form. The prepared powder was finally passed through a 125 μm

138   sieve for further microstructural analysis.

139

140   ICP-OES (PerkinElmer Optima DV2000) was employed to measure the chemical composition

141   of the reject brine before and after the reactions. The XRD (Bruker D8 Advance) diffractagrams

142   of the synthesized HMCs were recorded from 5° to 70° at 0.02°/step with a CuKα radiation at

143   40 kV and 40 mA. The morphology of the synthesized HMCs was investigated by a field

144   emission SEM (JSM-7600F). TG/DTA (PyrisDiamond 4000) of the synthesized HMCs was

145   operated at a heating rate of 10°C/min under the air flow condition.

146

147   **3. Results and Discussion**

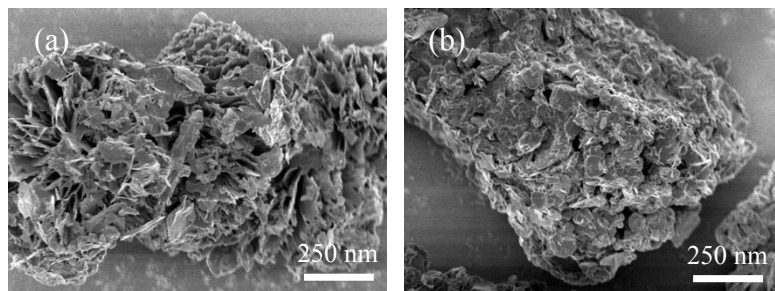148   **3.1. Characterization of HMCs synthesized from Mg(OH)$_2$ slurry**

149   **3.1.1. Effect of Mg(OH)$_2$/CO$_2$ molar ratio**
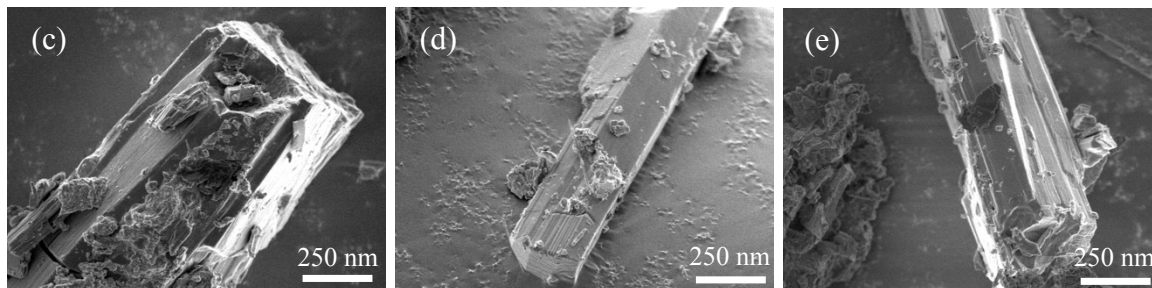
150   Figure 2 shows the FESEM images of all samples obtained under a constant pH of 8 while the

151   Mg(OH)$_2$:CO$_2$ molar ratio was varied between 1:1 and 1:7. The morphologies of the obtained

152   HMCs dramatically changed with the Mg(OH)$_2$:CO$_2$ molar ratio. For instance, the rosette-like

153   morphology observed when the Mg(OH)$_2$:CO$_2$ molar ratio was 1:1 (Fig. 2a), which was

154   eventually replaced by rod-like structures with smooth surfaces when this ratio gradually

155   increased to 1:6 (Figs. 2b-f). The boundaries of these rod-like carbonate phases became clearer

156   with an increase in the Mg(OH)$_2$:CO$_2$ molar ratio. A further increase in the Mg(OH)$_2$:CO$_2$

157   molar ratio to 1:7 revealed the formation of a layer of rosette-like flakes around the original

158   rod-like morphology, producing a "house of cards" texture on the surface [52], as shown in
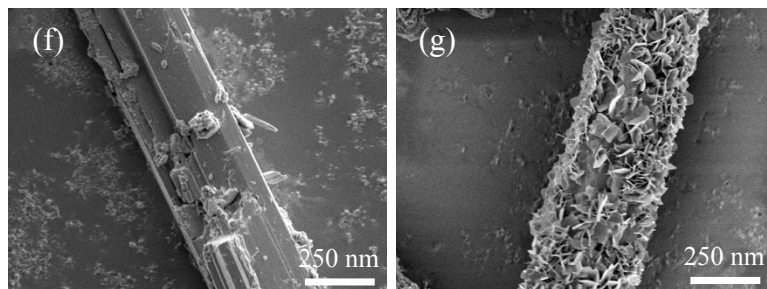
159   Fig. 2g.

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Figure 2 FESEM images of HMCs obtained under a pH of 8 at different $Mg(OH)_2:CO_2$ molar

ratios of (a) 1:1, (b) 1:2, (c) 1:3, (d) 1:4, (e) 1:5, (f) 1:6 and (g) 1:7

166

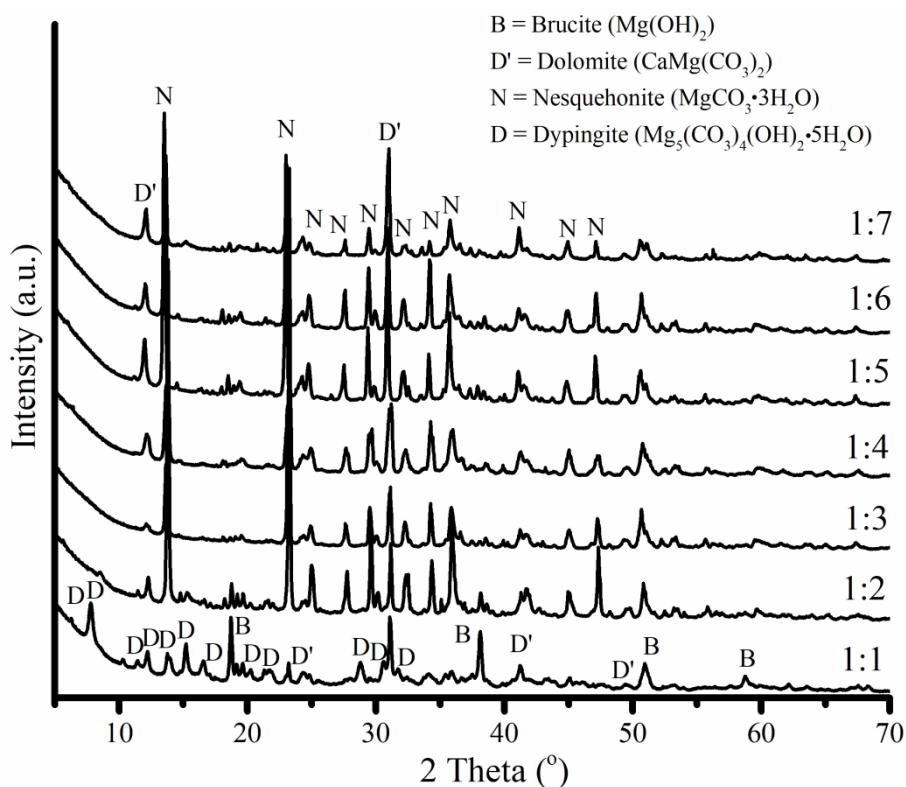Fig. 3 indicates the XRD diffractograms of the same set of samples obtained under a pH of 8

at different $Mg(OH)_2:CO_2$ molar ratios. The XRD patterns confirmed that the rosette- and rod-

like particles observes in Fig. 2 could be attributed to dypingite and nesquehonite, respectively.

These morphological observations were in line with the previous literature [18, 36, 41], where

the distinct morphologies of dypingite and nesquehonite were reported. At the $Mg(OH)_2:CO_2$

molar ratio of 1:1, the precipitates consisted of dypingite, uncarbonated brucite and dolomite

that was present as an impurity within $Mg(OH)_2$. An increase in the $Mg(OH)_2:CO_2$ molar ratio

to 1:2 revealed a reduction in the amount of uncarbonated brucite, resulting in the domination

of nesquehonite. These results corresponded well with the chemical composition of different

176 carbonate phases. Accordingly, the abundance of nesquehonite could be associated with the

177 availability of higher amounts of CO_2 introduced into the mix under higher $Mg(OH)_2$:CO_2

178 molar ratios. This is because nesquehonite ($MgCO_3 \cdot 3H_2O$) requires a higher Mg:CO_2 molar

179 ratio of 1:1 than dypingite ($Mg_5(CO_3)_4(OH)_2 \cdot 5(H_2O)$), which can theoretically form at a

180 corresponding ratio of 1:0.8. Further increase of the $Mg(OH)_2$:CO_2 molar ratio from 1:2 to 1:7

181 did not lead to significant changes in the XRD patterns, where nesquehonite continued to be

182 the dominate phase which is consistent with the SEM observation (Fig. 2).

183

184 The "house of cards" morphology is related to a dissolution-recrystallization self-assembly

185 growth mechanism when the dissolution rate of nesquehonite was faster than the precipitation

186 rate of hydromagnesite [52]. Thus, the formation of "house of cards" texture on the surface of

187 HMCs synthesized at $Mg(OH)_2$:CO_2 molar ratio of 1:7 (Fig. 2g) reveals that while

188 nesquehonite is still the dominating phase of the main body (Fig. 3), nesquehonite on the

189 surface was transformed into hydromagnesite phase due to elevated CO_2 concentration.

190

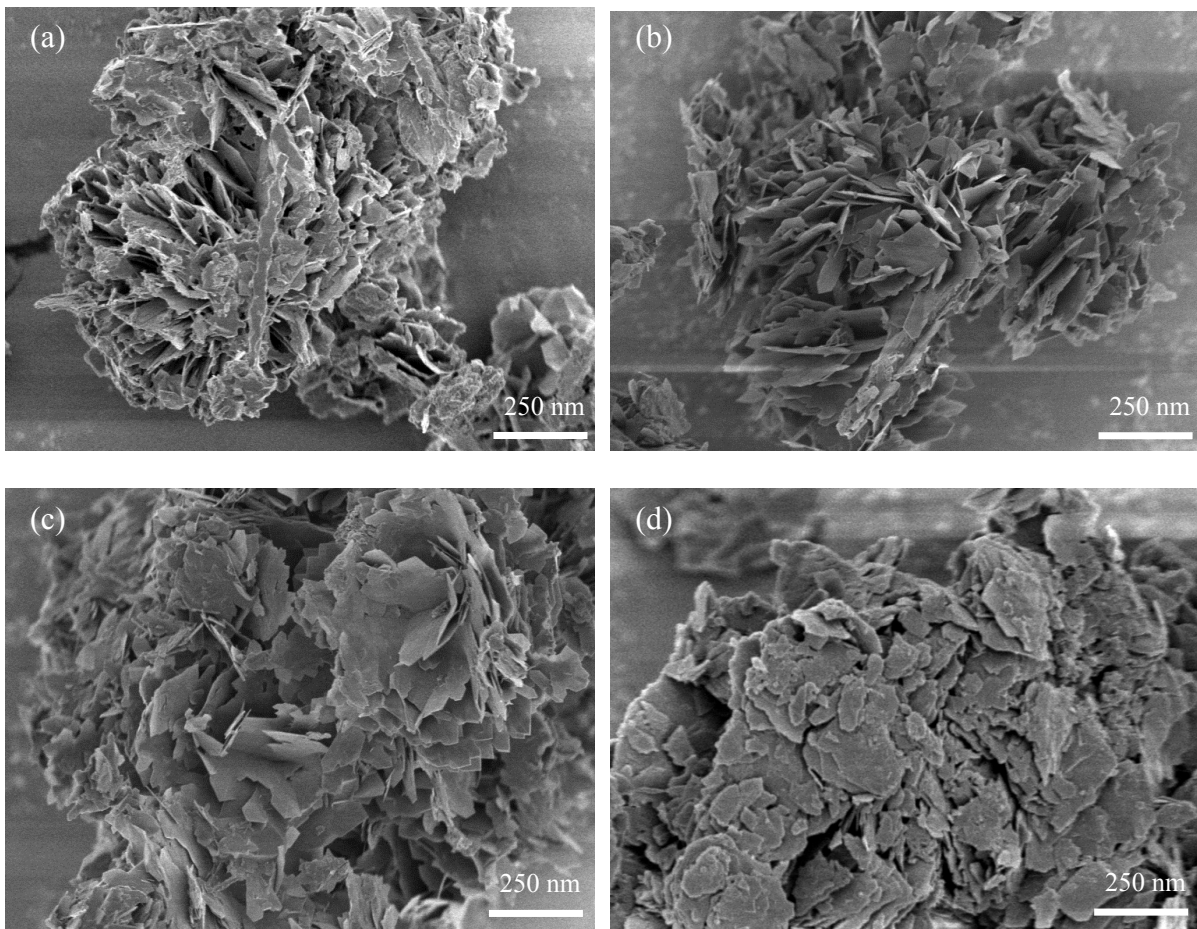Figure 3 XRD diffractograms of HMCs obtained under a pH of 8 at different $Mg(OH)_2$:CO_2 molar ratios

3.1.2. Effect of pH at $Mg(OH)_2$-to-CO_2 molar ratio of 1

Figure 4 illustrates the FESEM images of the samples obtained under different pH values ranging between 8 and 11, at a constant $Mg(OH)_2$:CO_2 molar ratio of 1:1. At the lower pH values of < 11, the obtained carbonates displayed rosette-like morphologies with an average dimension of ~2 μm, as shown in Figures 4(a)-(c). These rosette-like formations were confirmed to be dypingite, as shown by the XRD patterns presented in Figure 5. As the pH increased from 8 to 10, the intensity of the uncarbonated brucite peak revealed a decrease relative to the others, possibly indicting a reduction in the amount of brucite and an associated higher degree of carbonation at elevated pH levels. This increase in the carbonation degree could be associated with the higher CO_3^{2-}:HCO_3^- ratios in the prepared solutions at elevated pH levels. An increase in the pH led to higher concentrations of OH^-, therefore enabling the

206　conversion of HCO_3^- to CO_3^{2-}, which then reacted with Mg^{2+}, leading to the precipitation of

207　higher amounts of HMCs in the solution. Alternatively, a further increase in the pH from 11 to

208　14 lowered the carbonation degree of brucite, which was observed with its flake-like

209　morphology in Figure 4(d). These results were in line with the findings reported in previous

210　studies, where the optimal pH for the carbonation of brucite was shown to be around 9 [53].
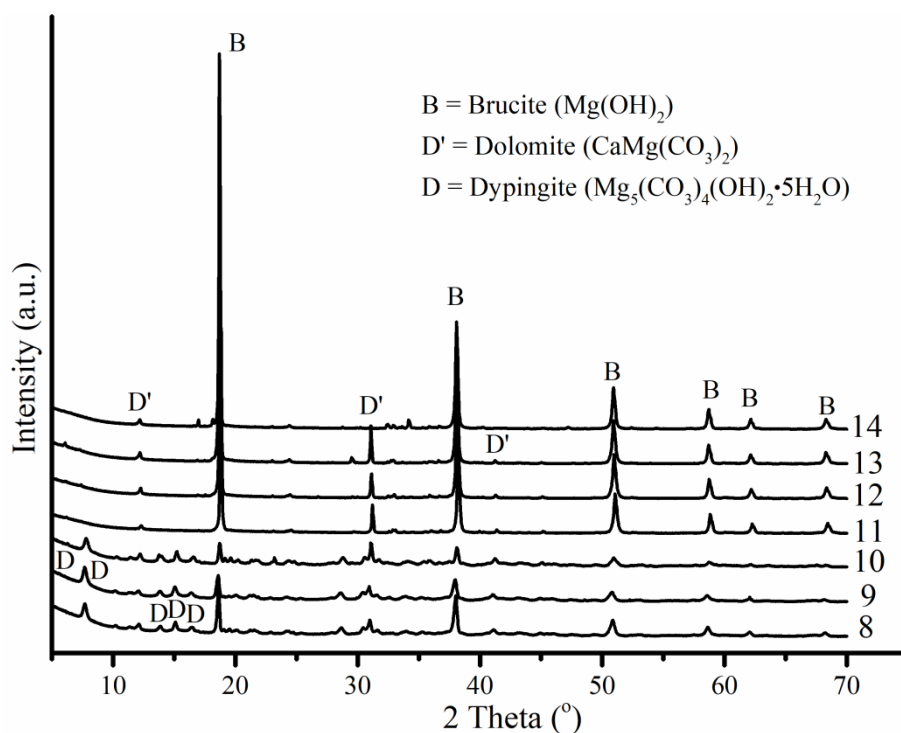
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213

214　Figure 4 FESEM images of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under

215　different pH values of (a) 8, (b) 9, (c) 10 and (d) 11

216

Figure 5 XRD diffractograms of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1

under different pH values

3.1.3 Effect of pH at $Mg(OH)_2$:CO_2 molar ratio of 1:2

Figure 6 displays the morphologies of HMCs obtained under different pH values at the $Mg(OH)_2$:CO_2 molar ratio of 1:2. Different from HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1, where the presence of dypingite with a rosette-like morphology dominated regardless of the pH value; HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:2 clearly demonstrated a different morphology. Instead of the previously observed rosette-like plates, a rod-like structure presenting the "house of cards" texture was seen in samples obtained under pH values of 8 and 9 (Figs. 6a-b). An increase in the pH from 8 onwards resulted in the distortion of the originally clear borders of the nesquehonite crystals, whose shape transformed from the rod-like structure to a cluster of flakes forming on top. This change was mostly obvious at pH values of 10 and 11 (Figures 6(c) and (d)), which revealed the formation of flake-
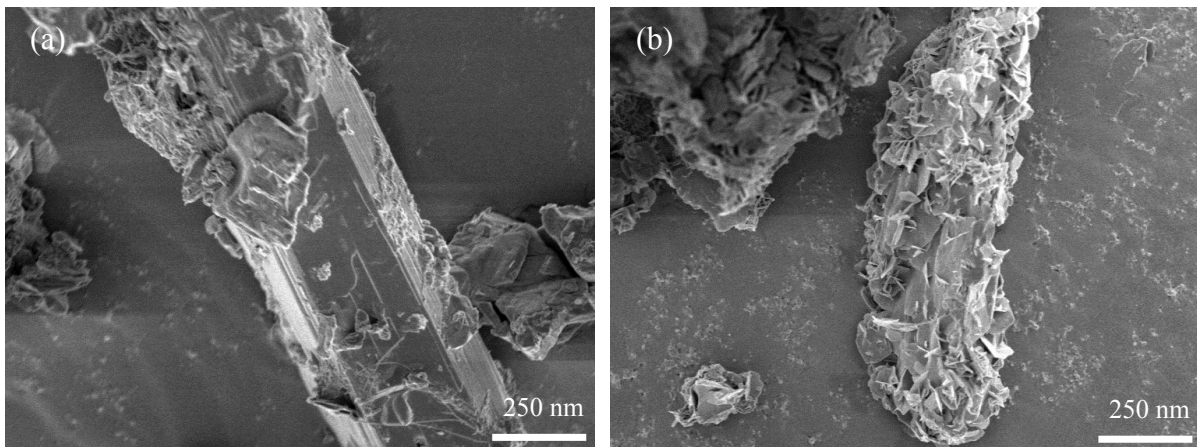
232   like clusters with clearly defined boundaries at a pH of 11.

233

234   The formation of nesquehonite under pH value of 11 was confirmed by the XRD patterns

235   shown in Figure 7. In line with the findings obtained under the $Mg(OH)_2$:CO_2 molar ratio of

236   1:1 as revealed in Figure 5, an increase in the pH to 10 and above resulted in the formation of

237   uncarbonated brucite where a flake-like morphology was observed. This "house of cards"

238   texture observed within the prepared samples was attributed to the dissolution-

239   recrystallization-self-assembly growth mechanism as explained in the aforementioned text.

240   The elevated pH used in the experiments conducted in this study increased the solubility of

241   CO_2 in the solution. This has led to a dissolution of the surface of nesquehonite and served as

242   the nucleation points for further hydromagnesite/dypingite plates growing with excessive CO_2

243   at the surface.

244

245

Figure 6 Typical FESEM images of HMCs obtained at a Mg(OH)$_2$:CO$_2$ molar ratio of 1:2

under different pH values of (a) 8, (b) 9, (c) 10 and (d) 11

Figure 7 XRD diffractograms of HMCs obtained at the Mg(OH)$_2$:CO$_2$ molar ratio of 1:2

under different pH values

3.2. Comparison of HMCs synthesized from different source

This section aims to provide a comparison of HMCs obtained via reject brine to those of

256   chemical $Mg(OH)_2$ slurry, whose detailed characterization was presented earlier in Section 3.1.

257   The findings presented here aim to use reject brine for the long-term storage of anthropogenic

258   CO_2.

259

260   **3.2.1. Microstructure of HMCs**

261   Figure 8 provides a comparison of the morphologies of HMCs synthesized from $Mg(OH)_2$

262   slurry and reject brine under a constant pH and $Mg(OH)_2$:CO_2 molar ratio of 8 and 1:1,

263   respectively. As shown in Figure 8(a), the carbonate crystals obtained via the use of $Mg(OH)_2$

264   slurry led to a rosette-like morphology. Alternatively, the carbonation of reject brine led to the

265   formation of a needle-like morphology with clear boundaries, as seen in Figure 8(b). The

266   compositions of these rosette- and needle-like particles were confirmed to be nesquehonite and

267   dypingite by XRD patterns revealed in Figure 9, respectively. The formation of different Mg-

268   carbonate phases via the two sources could be associated with the relatively higher reactivity

269   of $Mg(OH)_2$ prepared from reject brine when compared to that of $Mg(OH)_2$ slurry (i.e. with a

270   specific surface area of 7.4 vs. 4.8 m^2/g as tested by BET analysis). The carbonation of

271    $Mg(OH)_2$ with a higher reactivity could have capture more CO_2 and enabled the formation of

272   nesquehonite as opposed to dypingite since nesquehonite requires a higher Mg:CO_2 molar ratio

273   as explained in the aforementioned text. This difference in the reactivity of the two samples

274   was also reflected by the absence of the residual brucite peaks in reject brine, as opposed to the

275   clearly defined uncarbonated brucite peaks observed in the $Mg(OH)_2$ slurry, as seen in Figure
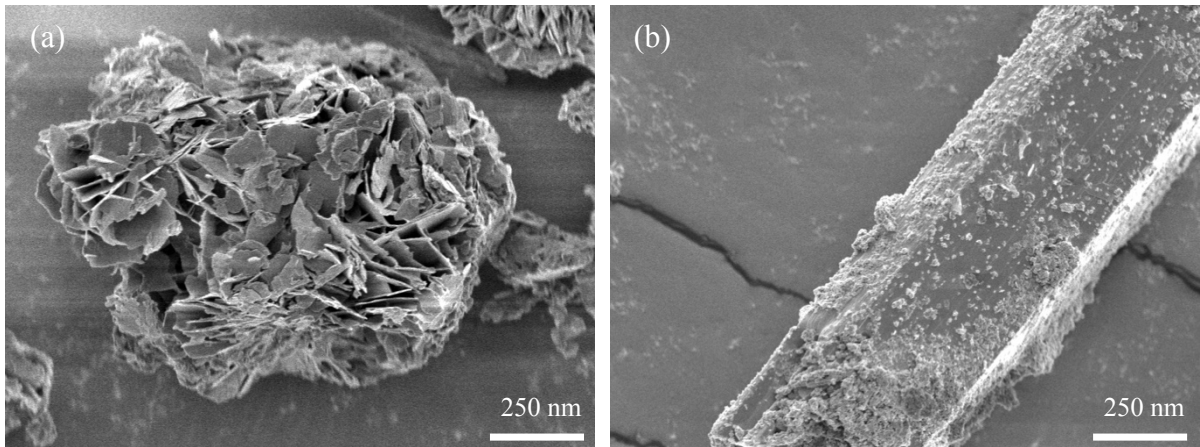
276   9.
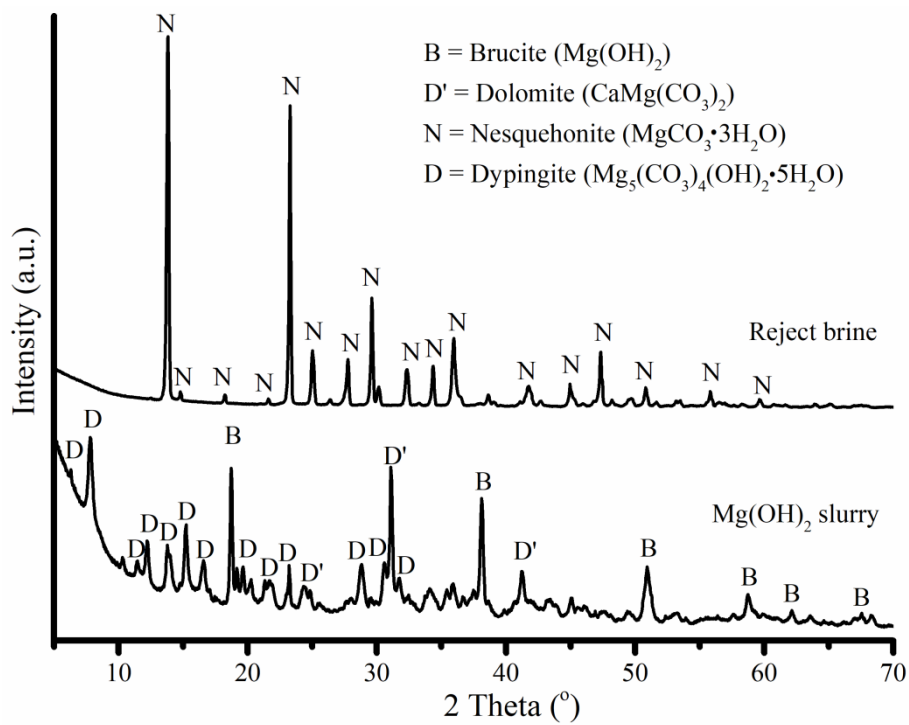
277

Figure 8 SEM images of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8, showing (a) $Mg(OH)_2$ slurry and (b) reject brine

Figure 9 XRD diffractograms of HMCs obtained from $Mg(OH)_2$ slurry and reject brine at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8

3.2.2. CO_2 capture and storage

287   The quantitative analysis of the chemical composition of HMCs and amount of CO_2 used in

288   their formation was carried out via TG/DTA and ICP-OES analyses. Figure 10 presents the

289   TG/DTA graphs of HMCs obtained from $Mg(OH)_2$ slurry and reject brine at the $Mg(OH)_2$:CO_2

290   molar ratio of 1:1 and under a pH 8. Both systems demonstrated a similar trend with three

291   stages of mass loss, which corresponded well with previous studies [36, 54, 55]. Firstly, the

292   dehydration of HMCs took place at ~100-250 °C, resulting in the loss of H_2O. The second mass

293   loss occurred between 250 and 550 °C, which was because of the decomposition of

294   uncarbonated $Mg(OH)_2$ into MgO as well as the decarbonation of HMCs, resulting in a loss of

295   H_2O and CO_2. The final mass loss observed between 550 and 700 °C was due to the

296   decomposition of dolomite and calcite, respectively, which was present as an impurity in the

297   two systems (i.e. in the form of dolomite in $Mg(OH)_2$ slurry and calcite in reject brine).

298

300   Figure 10 Typical TG/DTA curves of HMCs obtained from $Mg(OH)_2$ slurry and reject brine

301   at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8

302

303　The chemical composition of the final product was determined via a combination of the results

304　generated by TG/DTA and ICP-OES, where the recovery rate of Mg^{2+} was measured via ICP-

305　OES and TG/DTA was used to determine the quantity of each phase derived from XRD results.

306　Table 2 revealed the composition of the precipitate synthesized via the carbonation of $Mg(OH)_2$

307　slurry to be composed of 62.1% dypingite and 30.5% uncarbonated brucite. On the other hand,

308　the carbonation of reject brine led to a precipitate composed of 93.1% nesquehonite and 3.7%

309　uncarbonated brucite, as detailed in Table 2. These results were used in the calculation of the

310　percentage of captured CO_2, which was derived by measuring the mass of CO_2 in the HMCs

311　(i.e. dypingite/nesquehonite) divided by the initial input of CO_2 degassed to the system at the

312　$Mg(OH)_2$:CO_2 molar ratio of 1:1. The outcome of these analyses indicated that 43.7% CO_2

313　could be sequestered in the form of dypingite via the defusing CO_2 into $Mg(OH)_2$ slurry, while

314　the corresponding ratio of CO_2 sequestered in reject brine was calculated to be 82.6%. The

315　higher efficiency of CO_2 sequestration achieved via the use of reject brine was associated with

316　the increased reactivity of $Mg(OH)_2$ synthesized from reject brine. Details of the chemical

317　composition and the percentage of captured CO_2 were included in the Appendix section.

318

319　Table 2 Chemical composition of HMCs synthesized from $Mg(OH)_2$ slurry and reject brine at

320　　　　　　　　the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8

$Mg(OH)_2$ slurry	Mass (g)	Mass (%)	Reject brine	Mass (g)	Mass (%)
Dypingite	0.6	62.1	Nesquehonite	1.51	93.1
Brucite	0.07	30.5	Brucite	0.06	3.7
Dolomite	0.3	7.4	Calcite	0.05	3.2
CO_2 captured (%)	43.7		CO_2 captured (%)	82.6	

321

322　According to the Paris Agreement, it aims to reduce GHG emissions by 20% (*i.e.,* 7 gigatonnes

323　CO_2 emission reduction as annual anthropogenic CO_2 emissions is about 35 gigatonnes [56])

324　in order to hold the increase in the global average temperature to below 2°C above pre-

industrial levels [56]. According to the International Desalination Association, the global daily production of desalinated water generated by 18,426 desalination plants worldwide exceeds 86.8 million m^3 [47]. It is estimated that an equivalent amount of reject brine is generated [47]. The concentration of Mg^{2+} in reject brine is around 1700 ppm, and thus around 54 million tonnes of Mg could be recovered every year. With a carbon capture and storage rate of 82.6% in the current study, around 45 million tonnes of CO_2 can be sequestrated annually. The suggested methodology thus contributes to around 1% of the required CO_2 emission reduction (*i.e.,* 7 gigatonnes) aimed in the Paris Agreement.

4. Summary and Conclusions

This study presented the influences of key parameters including the $Mg(OH)_2$:CO_2 molar ratio, pH, and $Mg(OH)_2$ source on the synthesis of HMCs through the carbonation of $Mg(OH)_2$ slurry. The resulting HMCs were characterized via a combination of techniques including XRD, FESEM, and TG/DTA. Main conclusions drawn from this study are summarized below.

- The carbonation of $Mg(OH)_2$ slurry under the elevated of $Mg(OH)_2$:CO_2 molar ratio resulted in the transformation of dypingite to nesquehonite.
- Increasing the pH from 8 to 10 was found to promote the carbonation process of $Mg(OH)_2$, resulting in a higher carbonation degree.
- A specific "house of cards" texture, involving the formation of rosette-like dypingite flakes on the surface of nesquehonite needles, was discovered under elevated pH and $Mg(OH)_2$:CO_2 ratios conditions. The formation of this structure was associated with to a dissolution-recrystallization-self-assembly growth mechanism as nesquehonite was seen as a precursor for the further nucleation and seeding of hydromagnesite/dypingite on the surface.

- Carbonation of $Mg(OH)_2$ slurry synthesized from reject brine led to high yield, high purity, and high carbonation degree (82.6%) HMCs. Reject brine shows high potential to be used for capture and long-term storage of CO_2 in the form of HMCs.

The use of compressed commercial CO_2 in the research was to provide a pure source to evaluate CO_2 sequestration rate of the synthesized $Mg(OH)_2$ from desalination reject brine, which served as a model cast study. In the 'real world' case, different CO_2 sources and collecting methods such as CO_2 generated from factories, coal burning power plants, and municipal solid waste incineration plants, may be used. However, further study is necessary to evaluate CO_2 sequestration efficiency of the synthesized $Mg(OH)_2$ with different CO_2 sources and collecting methods. Furthermore, it is necessary to evaluation of the mass, energy, reagents, wastes that come into play in the global process from the life cycle and life cycle cost viewpoint of HMCs synthesized from reject brine.

Acknowledgement

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521

522

Appendix – Chemical composition of HMCs and CO₂ captured percentage

HMCs synthesized from $Mg(OH)_2$ slurry

Initially, 0.82 g $Mg(OH)_2$ is added into 200 ml distil water at a $Mg(OH)_2/CO_2$ molar ratio of 1. Assuming a purity of 92% for $Mg(OH)_2$, the remaining 8% impurity is dolomite which does not react or dissolve in the solution. The final precipitates consist of uncarbonated brucite, dypingite and dolomite as supported by XRD results (Figures 5 and 7), which after calcination decompose into MgO, MgO, and CaO·MgO, respectively. Let x and y denote the weights of the uncarbonated brucite and dypingite, respectively. The weight of dolomite is 0.066 g calculated based on 8% of the initial sample weight (i.e., 0.82 g). Residues after the TG/DTA test is 50.6% (Figure 10). Based on the given information, the following equation can be established.

$$\frac{\frac{40}{58}x + \frac{40}{96.8}y + 0.82 \times 0.08 \times \frac{96}{184}}{x + y + 0.82 \times 0.08} = 0.506 \tag{A1}$$

Furthermore, the concentration of Mg^{2+} in the residue was measured to be 308.4 ppm (0.308 g/L) as shown in Table 3. Thus,

$$\frac{\frac{24}{58}x + \frac{24}{96.8}y}{0.2} = \frac{\frac{24}{58} \times 0.82 \times 0.92}{0.2} - 0.308$$
$$\text{(A2)}$$

By solving the two Eqns. (A1) and (A2), the weights of uncarbonated brucite and dypingite are calculated to be 0.27 g and 0.55 g, respectively. The sum of the weights of uncarbonated brucite, dolomite and dypingite was calculated to be 0.89 g, which was higher than 0.46 g of the weighted precipitate as shown in Table 3. This was mainly due to the weight losses during the process of separating the samples from the solution and grinding. The captured CO_2 percentage was calculated by measuring the weight of CO_2 in the HMCs (dypingite) divided by the initial

547   input of CO_2 degassed to the system at the $Mg(OH)_2/CO_2$ molar ratio of 1, which was calculated

548   to be 43.7% in consideration of 8% impurity.

549

550   Table A1 Concentration of Mg^{2+} in the residue and final weight of HMCs obtained from

551    $Mg(OH)_2$ slurry and reject brine

	Mg^{2+} in the residue solution (ppm)	Weight of solids (g)
$Mg(OH)_2$ slurry	309.9 ± 5.2	0.46 ± 0.20
Reject brine	161.5 ± 9.1	1.24 ± 0.14

552

553   <u>HMCs synthesized from reject brine</u>

554   The same principal applies to HMCs synthesized from the reject brine. The $Mg(OH)_2$ sample

555   precipitated from reject brine in the first step contains 6.3% calcite as the impurity which has

556   been detailed in [50]. The XRD result confirms that HMCs after carbonation of $Mg(OH)_2$

557   consisted of nesquehonite, uncarboned brucite and calcite, which after calcination decompose

558   into **MgO, MgO and CaO**, respectively. Let x and y denoted the weights of uncarboned brucite

559   and nesquehonite, respectively. The weight of calcite is 0.052 g calculated based on 6.3% of

560   the initial sample weight (i.e., 0.82 g). Residues after the TG/DTA test is 31.3% (Figure 10).

561   Based on the given information, the following equation can be established.

562
$$\frac{\frac{40}{58}x + \frac{40}{138}y + 0.82 \times 0.063 \times \frac{56}{100}}{x + y + 0.82 \times 0.063} = 0.313 \qquad (A3)$$

563   Furthermore, the concentration of Mg^{2+} in the residue brine was measured to be 159.3 ppm

564   (0.159 g/L) as shown in Table 3. Thus,

565
$$\frac{\frac{24}{58}x + \frac{24}{138}y}{0.2} = \frac{\frac{24}{58} \times 0.82 \times 0.937}{0.2} - 0.159 \qquad (A4)$$

566

567   By solving the two Eqns. (A1) and (A2), the weights of uncarbonated brucite and nesquehonite

568   are calculated to be 0.06 g and 1.51 g, respectively. The captured CO_2 percentage was

569   calculated by measuring the weight of CO_2 in the HMCs (nesquehonite) divided by the initial

570   input of CO_2 degassed to the system at the $Mg(OH)_2/CO_2$ molar ratio of 1, which was calculated

571   to be 82.6% in consideration of 6.3% impurity. A successful sequestration of CO_2 into reject

572   brine as HMCs was therefore achieved to obtain an efficiency as high as 82.6%, which was

573   significantly improved compared to the $Mg(OH)_2$ slurry.

574

- Carbonation of $Mg(OH)_2$ slurry obtained from reject brine is studied

- $Mg(OH)_2$:CO_2 ratio and pH influence the type and amount of carbonates

- Formation of nesquehonite is observed at high $Mg(OH)_2$:CO_2 ratios and pH

- A "house of cards" texture is formed with an increase in pH and $Mg(OH)_2$:CO_2 ratio

- Carbonation of $Mg(OH)_2$ slurry obtained from reject brine has a CCS efficiency of 82.6%

1   **Microstructure and carbon storage capacity of hydrated magnesium carbonates**

2   **synthesized from different sources and conditions**

3

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5

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10

11 **Abstract**

12 Recently, the mineral carbonation via the reaction of CO_2 with saline aquafers received much

13 attention as one of the most promising ways for geologic CO_2 storage. This paper reports

14 microstructure and carbon storage capacity of hydrated magnesium carbonates (HMCs)

15 synthesized from different sources, *i.e.,* reject brine and commercial $Mg(OH)_2$ slurry, and under

16 different conditions, *i.e.,* pH (8-14) and $Mg(OH)_2$:CO_2 molar ratio (1:1-1:7). Results show that

17 dypingite ($Mg_5(CO_3)_4(OH)_2 \cdot 5H_2O$) is the main phase forming at lower $Mg(OH)_2$:CO_2 ratios.

18 An increase in the $Mg(OH)_2$:CO_2 ratio and/or pH leads to the precipitation of nesquehonite

19 ($MgCO_3 \cdot 3H_2O$). A unique "house of cards" texture, involving formation of the rosette-like

20 dypingite flakes on the surface of nesquehonite needles, is discovered under elevated pH and

21 $Mg(OH)_2$:CO_2 ratios. HMCs synthesized from reject brine exhibit a much higher carbon

22 storage capacity of 82.6% than that produced from the commercial $Mg(OH)_2$ slurry (43.7%).

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23 Findings from this study advance understanding of mineral recovery from reject brine and the

24 capture and long-term storage of CO_2 in the form of HMCs.

25

28

29 **1. Introduction**

30 Fossil fuels have been the world's primary energy source, providing over 85% of the energy

31 demands worldwide [1-6]. However, nearly 83% of the anthropogenic greenhouse gas (GHG)

32 emissions are coming from combustion and non-fuel uses of fossil fuels [7]. CO_2, which is the

33 main GHG, has caused most of the global warming since it has the highest positive radiative

34 forcing and far more abundant in the atmosphere than other heat-trapping gases [8, 9]. The

35 concentration of CO_2 in the atmosphere has increased ~30% from 325 parts per million (ppm)

36 at the beginning of the industrial era in 1970 to 409.7 ppm in May 2017 measured in Mauna

37 Loa Observatory. Consequently, much attention has been drawn to the carbon management

38 [10-12]. Carbon capture and storage (CCS) provides a feasible way to reduce the build-up of

39 CO_2 in the atmosphere [13, 14]. CCS concept covers broad fields such as ocean, terrestrial,

40 geological, biological and chemical approaches to store CO_2 gas in the long term [15-18],

41 among which mineral carbonation via the reaction of CO_2 with saline aquafers is one of the

42 most promising geologic CO_2 storage options [19-23]. Magnesium-based minerals have

43 attracted great attentions worldwide as they show the potential to sequestrate anthropogenic

44 CO_2 to counterpart the global warming [24-27]. Furthermore, reactive magnesia (MgO) cement

45 has been studied as a potential alternative to the Portland cement due to its ability to sequester

46 significant amount of carbon dioxide (CO_2) and potential to be fully recycled [28-32].

47

48   Although the most thermodynamically stable carbonate for magnesium is in the anhydrous

49   form, *i.e.,* magnesite ($MgCO_3$) [33-35], formation of magnesite at the ambient condition is not

50   common. Instead, formation of hydrated magnesium carbonates (HMCs) prevails as Mg^{2+} ions

51   in the solution are highly hydrated [24]. HMCs are a class of magnesium compounds that form

52   in MgO-CO_2-H_2O systems, where the carbonation of magnesium systems generate a variety of

53   phases,   including   dypingite   ($Mg_5(CO_3)_4(OH)_2 \cdot 5H_2O$)   [36],   hydromagnesite

54   ($Mg_5(CO_3)_4(OH)_2 \cdot 4H_2O$) [37], and nesquehonite ($MgCO_3 \cdot 3H_2O$) [18, 38]. It has been reported

55   that inclusion of HMCs in reactive MgO cement enhances carbonation of the resulting binder
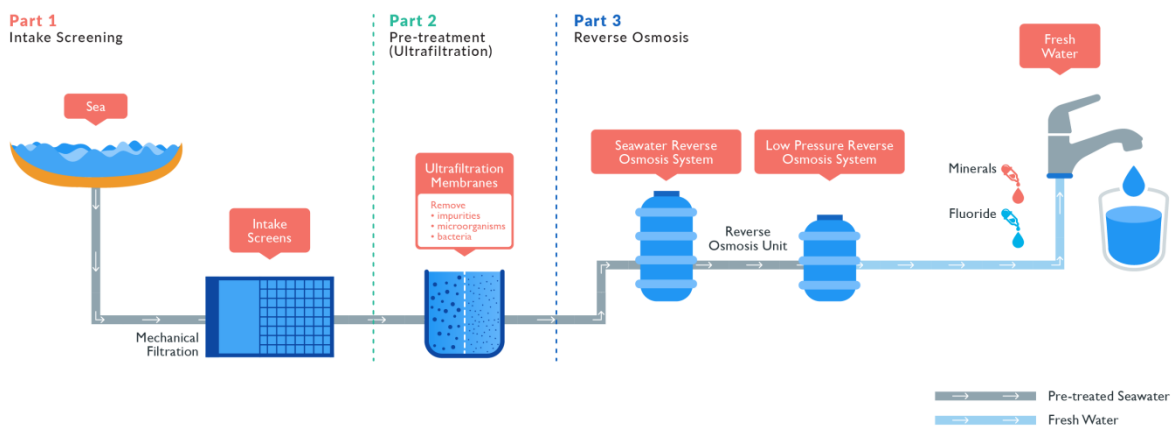
56   [39].

57

58   Theoretically, the formation of different phases of HMCs with different morphologies is

59   influenced by temperature, pH and CO_2 partial pressure [24, 40, 41]. Recent experimental

60   studies show that the precipitation of nesquehonite with needle-like morphology in an aqueous

61   solution occurs commonly under ambient conditions [15, 18, 38]. As the reaction temperatures

62   (333-368 K) and pH values increased, needle-like nesquehonite was replaced by

63   hydromagnesite with sheet-like morphology since nesquehonite is widely known to transform

64   to hydromagnesite at temperature above 50°C [40]. Hydromagnesite was reported to form at

65   120°C and P_{CO2} of 3 bar, which gradually transformed to magnesite within 5-15 hours.

66   However, a further increase of P_{CO2} to 100 bar at 120°C resulted in the direct precipitation of

67   magnesite [24]. The thermal behavior of the synthesized HMCs (*e.g.,* nesquehonite) has been

68   tested through real time in-situ X-ray diffraction (XRD), which indicated that nesquehonite and

69   dypingite remained thermal and structural stable up to 373 K and 435 K, respectively [36, 42].

70   Under the continued thermal treatment, nesquehonite transited into magnesite which was even

71   thermally stable up to 600 K, while dypingite transited into hydromagnesite at around 570 K,

72   assuring the long-term storage of CO_2. However, the influences of $Mg(OH)_2$/CO_2 molar ratios

73 and pH on the phases, morphology, and CCS efficiency of HMCs have yet been studied

74 systematically.

75

76 Desalination is a process that removes minerals from saline water. In coastal regions such as

77 Singapore where sources of fresh water are limited, desalination provides a feasible option to

78 produce fresh water. However, a high salty waste stream (*i.e.,* reject brine) would be generated

79 as much as the produced desalinated water at the end of the process [43]. Reject brine is a much

80 more complex media because chemicals are often added into the intake seawater (*e.g.,* to

81 precipitate the colloidal particles before running through the ultra-filtration) in the desalination

82 process (Fig. 1). Although many studies have investigated mineral trapping of CO_2 into saline

83 aquafers (*e.g.,* seawater, natural brine, or synthesized $MgCl_2$ solution) [18, 36, 42, 44-46], very

84 few has reported the use of reject brine as the CO_2 reservoir [47] and no study has proposed to

85 use reject brine as the $Mg(OH)_2$ source to synthesize HMCs.

86

87

88   Figure 1 Schematic illustration of the typical process in a reverse osmosis desalination plant

89   [48]

90

In this paper, influence of key parameters including $Mg(OH)_2$:CO_2 molar ratio, pH, and $Mg(OH)_2$ source on HMCs synthesis were investigated. The resulting HMCs were characterized by means of XRD, scanning electron microscopy (SEM), and thermogravimetric/differential thermal analysis (TG/DTA) to reveal phases, morphology, and CCS efficiency of HMCs synthesized under different conditions. In the following sections, materials and methodologies are presented first, followed by results presentation and discussion.

2. Materials and Methodology

2.1. Materials

Reject brine sample was collected from a local desalination plant which generates 318,500 m^3 desalinated water per day. The sample was filtrated through a 45 μm membrane filter to remove suspended solids before further analysis. An inductively coupled plasma-optical emission spectroscopy (ICP-OES) was used to analyze the chemical composition of reject brine in the current study (Table 1). Analytical grade $Mg(OH)_2$ (92% pure) and analytical grade sodium hydroxide (NaOH) with a purity of 97% were both purchased from VWR Pte Ltd in Singapore. Compressed CO_2 was purchased from Leeden National Oxygen Ltd in Singapore.

Table 1 Chemical composition of reject brine

Element	Cl	Na	SO$_4$	Mg	K	Ca	Sr	B	Si	Li	P	Al
Concentration (ppm)	65593	13580	4323	1718	846	471	14.6	3.8	3.7	0.3	0.2	0.1

2.2. Methodology

In the first approach, 0.82 g commercially available analytical grade $Mg(OH)_2$ was dissolved into 200 ml ultra-pure water to prepare the $Mg(OH)_2$ slurry. To study the influences of

`5

114    $Mg(OH)_2/CO_2$ molar ratios and pH on the microstructures of HMCs , three sets of experiments

115   were designed at controlled conditions, *i.e.,* $Mg(OH)_2:CO_2$ molar ratio (1:1-1:7) and pH (8-14).

116   CO_2 was sparged into the slurry at a rate of 100 ml/min at room temperature under pre-

117   determined conditions as follows,

118   a) pH = 8, $Mg(OH)_2:CO_2$ molar ratio = 1:1 to 1:7

119   b) pH = 8 to 11, $Mg(OH)_2:CO_2$ molar ratio = 1:1

120   c) pH = 8 to 11, $Mg(OH)_2:CO_2$ molar ratio = 1:2

121

122   In the second approach, $Mg(OH)_2$ was first synthesized from the reject brine via the addition

123   of NaOH at an optimized condition (*i.e.,* $NaOH/Mg^{2+}$ ratio of 2 at 25°C) determined from our

124   previous work [49-51], which results high yield and high purity $Mg(OH)_2$. After which, 0.82 g

125   of synthesized $Mg(OH)_2$ was mixed with 200 ml ultra-pure water. CO_2 was sparged into the

126   slurry at a rate of 100 ml/min at room temperature under controlled condition (*i.e.,* pH = 8,

127    $Mg(OH)_2:CO_2$ molar ratio = 1:1).

128

129   A pH/thermometer probe, inserted into reject brine, was used to continuously record the

130   temperature and pH during the experiment. A CO_2 flowmeter was used to monitor and record

131   the volume of CO_2 diffused into the slurry. Once the diffused CO_2 reached the pre-determined

132   value (*i.e.,* calculated based on the designed $Mg(OH)_2:CO_2$ molar ratio), the reaction was

133   terminated. 1M NaOH was added into the slurry continuously to maintain the pH to the

134   designed value since sparging CO_2 lowered the pH of slurry. HMCs were separated from the

135   solution by a centrifuge and washed three times using ultra-pure water. The washed samples

136   were fully dried in an oven at low temperature (*i.e.,* 40°C) to avoid any phase changes, before

137   being ground into powder form. The prepared powder was finally passed through a 125 μm

138   sieve for further microstructural analysis.

139

140   ICP-OES (PerkinElmer Optima DV2000) was employed to measure the chemical composition

141   of the reject brine before and after the reactions. The XRD (Bruker D8 Advance) diffractagrams

142   of the synthesized HMCs were recorded from 5° to 70° at 0.02°/step with a CuKα radiation at

143   40 kV and 40 mA. The morphology of the synthesized HMCs was investigated by a field

144   emission SEM (JSM-7600F). TG/DTA (PyrisDiamond 4000) of the synthesized HMCs was

145   operated at a heating rate of 10°C/min under the air flow condition.

146

147 **3. Results and Discussion**

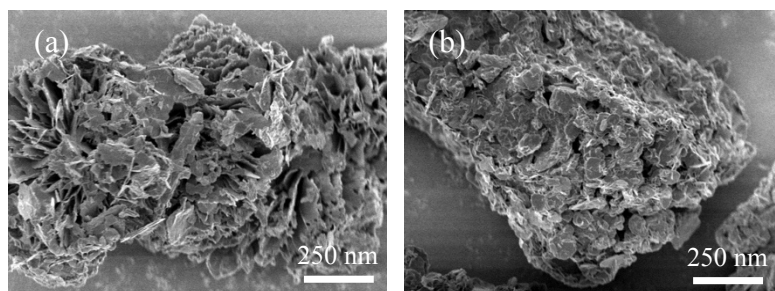148 **3.1. Characterization of HMCs synthesized from $Mg(OH)_2$ slurry**

149 **3.1.1. Effect of $Mg(OH)_2$/CO_2 molar ratio**

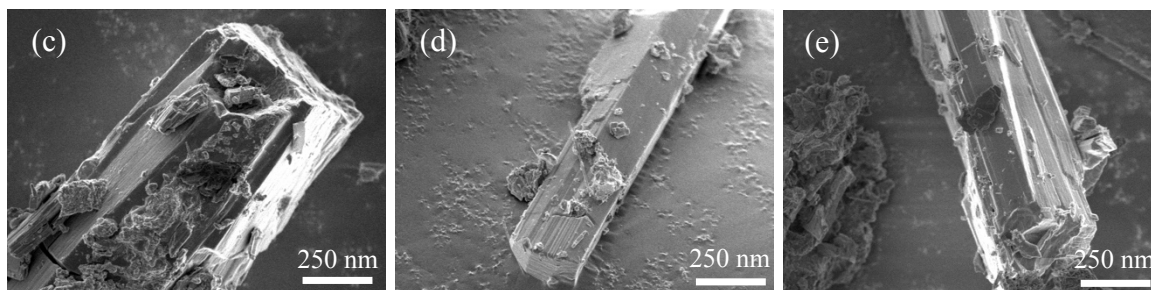150   Figure 2 shows the FESEM images of all samples obtained under a constant pH of 8 while the

151      $Mg(OH)_2$:CO_2 molar ratio was varied between 1:1 and 1:7. The morphologies of the obtained

152   HMCs dramatically changed with the $Mg(OH)_2$:CO_2 molar ratio. For instance, the rosette-like

153   morphology observed when the $Mg(OH)_2$:CO_2 molar ratio was 1:1 (Fig. 2a), which was

154   eventually replaced by rod-like structures with smooth surfaces when this ratio gradually

155   increased to 1:6 (Figs. 2b-f). The boundaries of these rod-like carbonate phases became clearer

156   with an increase in the $Mg(OH)_2$:CO_2 molar ratio. A further increase in the $Mg(OH)_2$:CO_2

157   molar ratio to 1:7 revealed the formation of a layer of rosette-like flakes around the original

158   rod-like morphology, producing a "house of cards" texture on the surface [52], as shown in
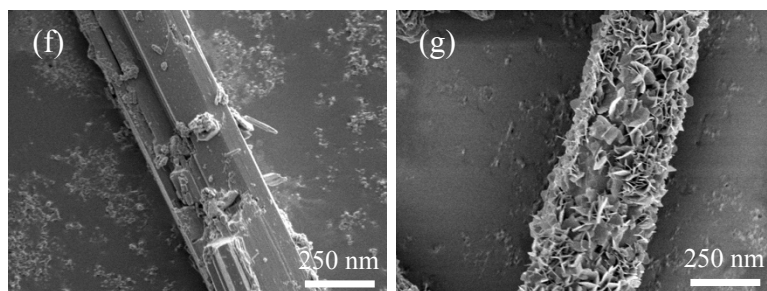
159   Fig. 2g.

160

161

162

163

164   Figure 2 FESEM images of HMCs obtained under a pH of 8 at different Mg(OH)$_2$:CO$_2$ molar

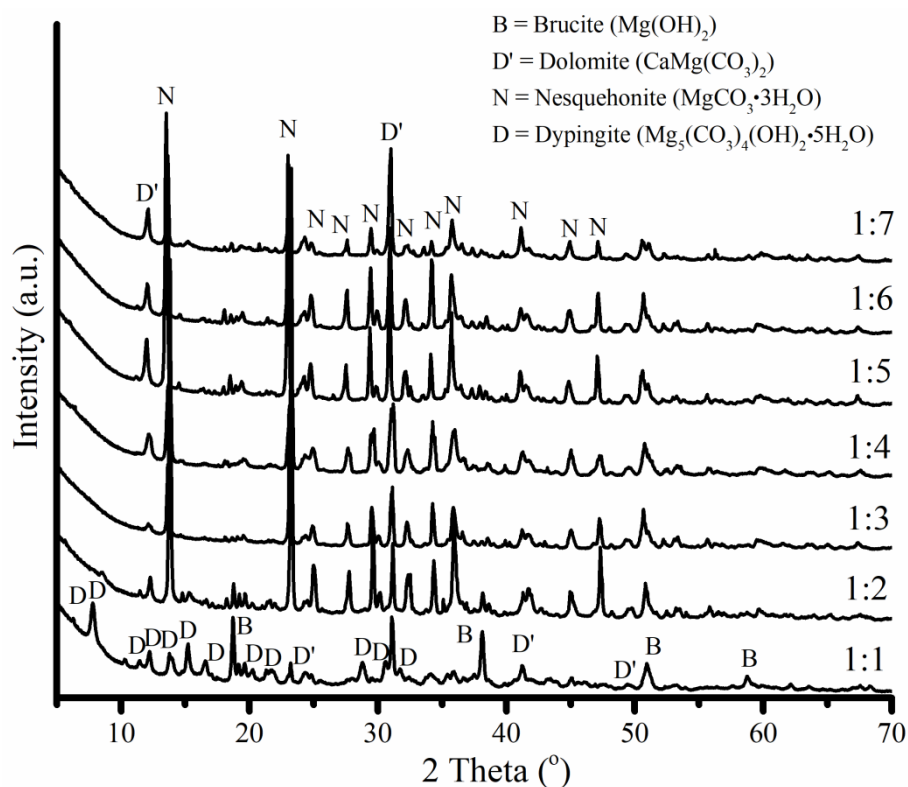165   ratios of (a) 1:1, (b) 1:2, (c) 1:3, (d) 1:4, (e) 1:5, (f) 1:6 and (g) 1:7

166

167   Fig. 3 indicates the XRD diffractograms of the same set of samples obtained under a pH of 8

168   at different Mg(OH)$_2$:CO$_2$ molar ratios. The XRD patterns confirmed that the rosette- and rod-

169   like particles observes in Fig. 2 could be attributed to dypingite and nesquehonite, respectively.

170   These morphological observations were in line with the previous literature [18, 36, 41], where

171   the distinct morphologies of dypingite and nesquehonite were reported. At the Mg(OH)$_2$:CO$_2$

172   molar ratio of 1:1, the precipitates consisted of dypingite, uncarbonated brucite and dolomite

173   that was present as an impurity within Mg(OH)$_2$. An increase in the Mg(OH)$_2$:CO$_2$ molar ratio

174   to 1:2 revealed a reduction in the amount of uncarbonated brucite, resulting in the domination

175   of nesquehonite. These results corresponded well with the chemical composition of different

176 carbonate phases. Accordingly, the abundance of nesquehonite could be associated with the

177 availability of higher amounts of CO_2 introduced into the mix under higher $Mg(OH)_2$:CO_2

178 molar ratios. This is because nesquehonite ($MgCO_3 \cdot 3H_2O$) requires a higher Mg:CO_2 molar

179 ratio of 1:1 than dypingite ($Mg_5(CO_3)_4(OH)_2 \cdot 5(H_2O)$), which can theoretically form at a

180 corresponding ratio of 1:0.8. Further increase of the $Mg(OH)_2$:CO_2 molar ratio from 1:2 to 1:7

181 did not lead to significant changes in the XRD patterns, where nesquehonite continued to be

182 the dominate phase which is consistent with the SEM observation (Fig. 2).

183

184 The "house of cards" morphology is related to a dissolution-recrystallization self-assembly

185 growth mechanism when the dissolution rate of nesquehonite was faster than the precipitation

186 rate of hydromagnesite [52]. Thus, the formation of "house of cards" texture on the surface of

187 HMCs synthesized at $Mg(OH)_2$:CO_2 molar ratio of 1:7 (Fig. 2g) reveals that while

188 nesquehonite is still the dominating phase of the main body (Fig. 3), nesquehonite on the

189 surface was transformed into hydromagnesite phase due to elevated CO_2 concentration.

190

Figure 3 XRD diffractograms of HMCs obtained under a pH of 8 at different $Mg(OH)_2$:CO_2 molar ratios

3.1.2. Effect of pH at $Mg(OH)_2$-to-CO_2 molar ratio of 1

Figure 4 illustrates the FESEM images of the samples obtained under different pH values ranging between 8 and 11, at a constant $Mg(OH)_2$:CO_2 molar ratio of 1:1. At the lower pH values of < 11, the obtained carbonates displayed rosette-like morphologies with an average dimension of ~2 µm, as shown in Figures 4(a)-(c). These rosette-like formations were confirmed to be dypingite, as shown by the XRD patterns presented in Figure 5. As the pH increased from 8 to 10, the intensity of the uncarbonated brucite peak revealed a decrease relative to the others, possibly indicting a reduction in the amount of brucite and an associated higher degree of carbonation at elevated pH levels. This increase in the carbonation degree could be associated with the higher CO_3^{2-}:HCO_3^- ratios in the prepared solutions at elevated pH levels. An increase in the pH led to higher concentrations of OH^-, therefore enabling the

206   conversion of HCO_3^- to CO_3^{2-}, which then reacted with Mg^{2+}, leading to the precipitation of

207   higher amounts of HMCs in the solution. Alternatively, a further increase in the pH from 11 to

208   14 lowered the carbonation degree of brucite, which was observed with its flake-like

209   morphology in Figure 4(d). These results were in line with the findings reported in previous

210   studies, where the optimal pH for the carbonation of brucite was shown to be around 9 [53].
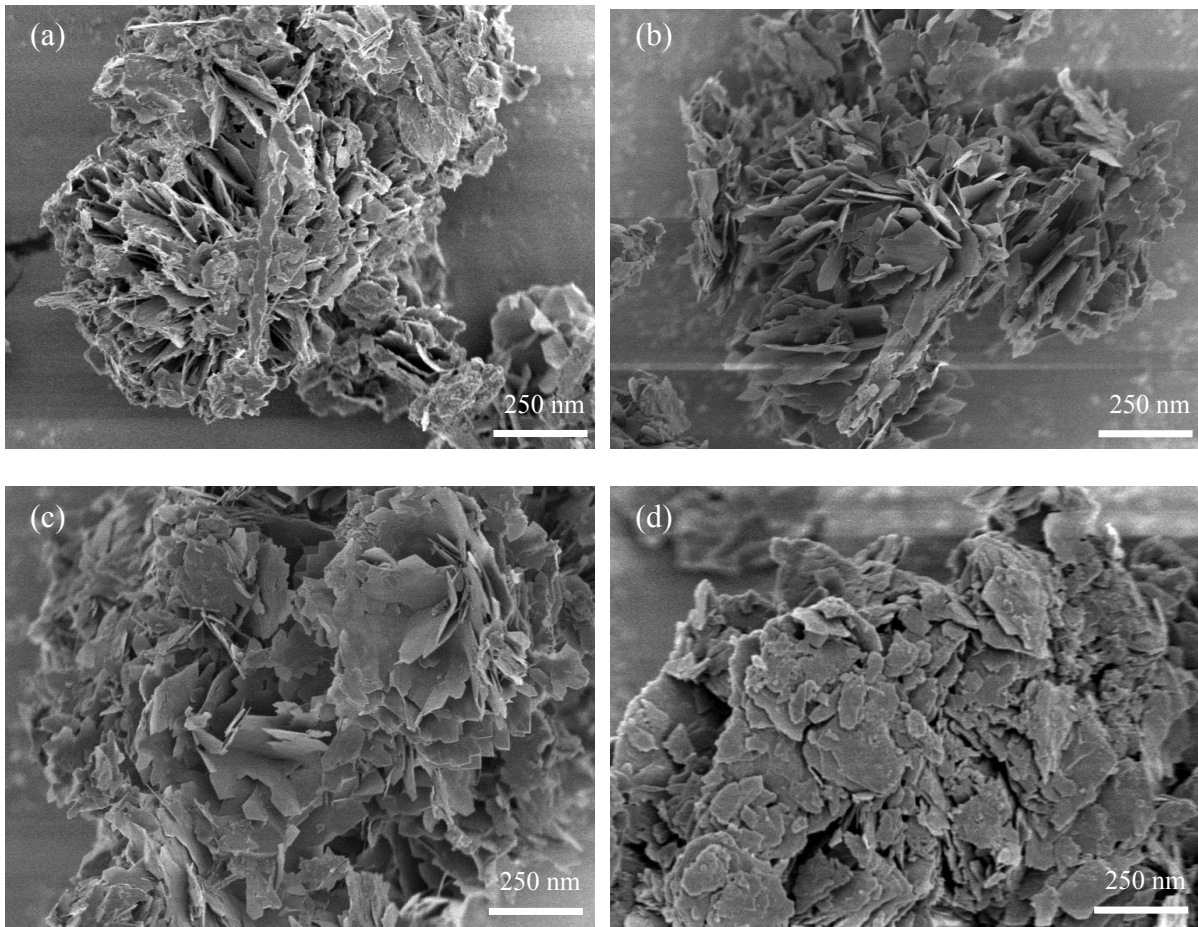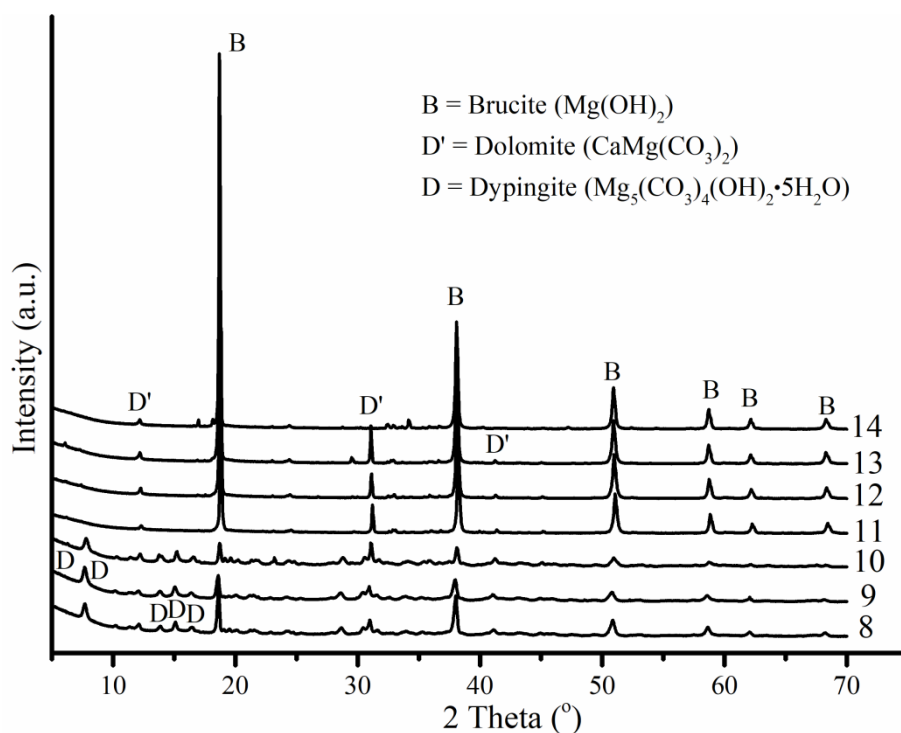
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213

214   Figure 4 FESEM images of HMCs obtained at the $Mg(OH)_2:CO_2$ molar ratio of 1:1 under

215   different pH values of (a) 8, (b) 9, (c) 10 and (d) 11

216

B = Brucite $(Mg(OH)_2)$
D' = Dolomite $(CaMg(CO_3)_2)$
D = Dypingite $(Mg_5(CO_3)_4(OH)_2 \cdot 5H_2O)$

Figure 5 XRD diffractograms of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1

under different pH values

3.1.3 Effect of pH at $Mg(OH)_2$:CO_2 molar ratio of 1:2

Figure 6 displays the morphologies of HMCs obtained under different pH values at the $Mg(OH)_2$:CO_2 molar ratio of 1:2. Different from HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1, where the presence of dypingite with a rosette-like morphology dominated regardless of the pH value; HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:2 clearly demonstrated a different morphology. Instead of the previously observed rosette-like plates, a rod-like structure presenting the "house of cards" texture was seen in samples obtained under pH values of 8 and 9 (Figs. 6a-b). An increase in the pH from 8 onwards resulted in the distortion of the originally clear borders of the nesquehonite crystals, whose shape transformed from the rod-like structure to a cluster of flakes forming on top. This change was mostly obvious at pH values of 10 and 11 (Figures 6(c) and (d)), which revealed the formation of flake-
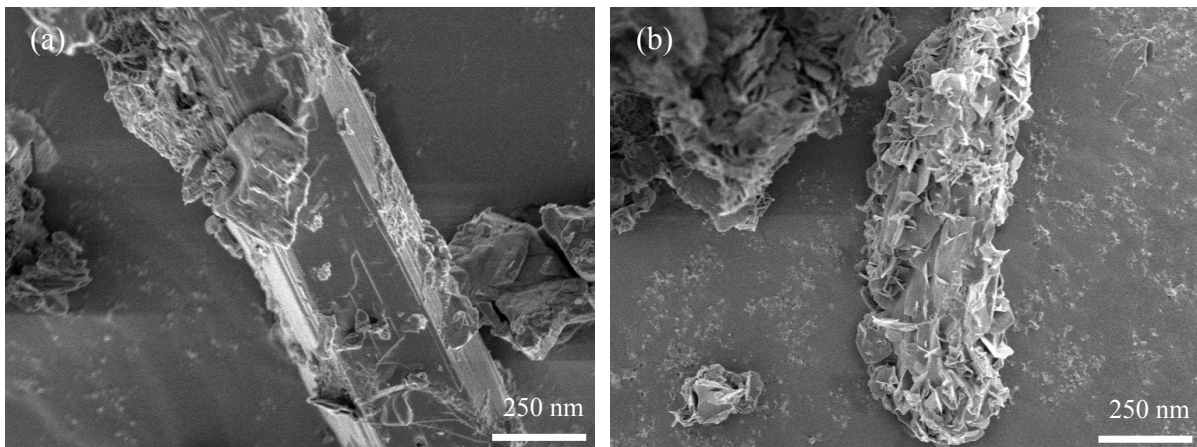
`12

232   like clusters with clearly defined boundaries at a pH of 11.

233

234   The formation of nesquehonite under pH value of 11 was confirmed by the XRD patterns

235   shown in Figure 7. In line with the findings obtained under the $Mg(OH)_2:CO_2$ molar ratio of

236   1:1 as revealed in Figure 5, an increase in the pH to 10 and above resulted in the formation of

237   uncarbonated brucite where a flake-like morphology was observed. This "house of cards"

238   texture observed within the prepared samples was attributed to the dissolution-

239   recrystallization-self-assembly growth mechanism as explained in the aforementioned text.

240   The elevated pH used in the experiments conducted in this study increased the solubility of

241   CO_2 in the solution. This has led to a dissolution of the surface of nesquehonite and served as

242   the nucleation points for further hydromagnesite/dypingite plates growing with excessive CO_2

243   at the surface.

244

245

`13

Figure 6 Typical FESEM images of HMCs obtained at a $Mg(OH)_2$:CO_2 molar ratio of 1:2

under different pH values of (a) 8, (b) 9, (c) 10 and (d) 11

Figure 7 XRD diffractograms of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:2

under different pH values

3.2. Comparison of HMCs synthesized from different source

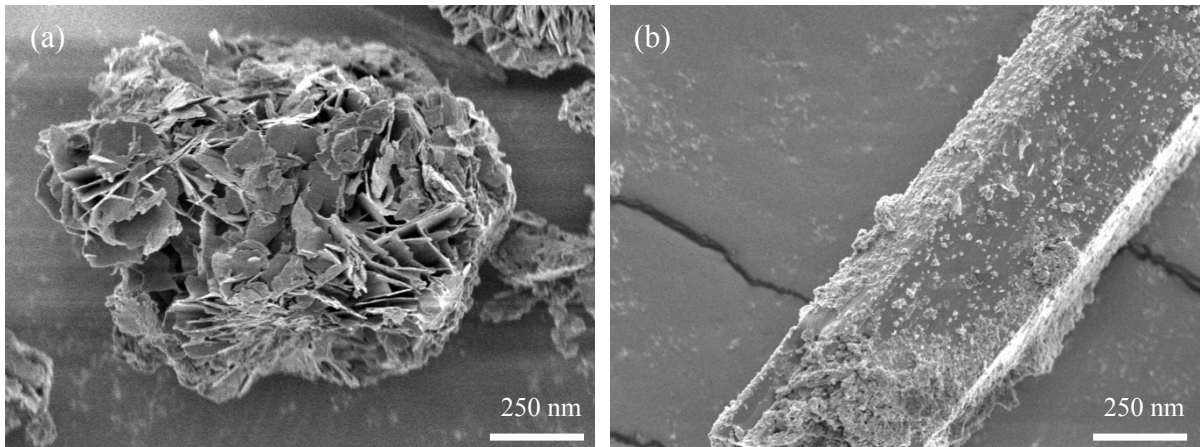This section aims to provide a comparison of HMCs obtained via reject brine to those of

256 chemical $Mg(OH)_2$ slurry, whose detailed characterization was presented earlier in Section 3.1.

257 The findings presented here aim to use reject brine for the long-term storage of anthropogenic

258 CO_2.

259

260 **3.2.1. Microstructure of HMCs**

261 Figure 8 provides a comparison of the morphologies of HMCs synthesized from $Mg(OH)_2$

262 slurry and reject brine under a constant pH and $Mg(OH)_2$:CO_2 molar ratio of 8 and 1:1,

263 respectively. As shown in Figure 8(a), the carbonate crystals obtained via the use of $Mg(OH)_2$

264 slurry led to a rosette-like morphology. Alternatively, the carbonation of reject brine led to the

265 formation of a needle-like morphology with clear boundaries, as seen in Figure 8(b). The

266 compositions of these rosette- and needle-like particles were confirmed to be nesquehonite and

267 dypingite by XRD patterns revealed in Figure 9, respectively. The formation of different Mg-

268 carbonate phases via the two sources could be associated with the relatively higher reactivity

269 of $Mg(OH)_2$ prepared from reject brine when compared to that of $Mg(OH)_2$ slurry (i.e. with a

270 specific surface area of 7.4 vs. 4.8 m^2/g as tested by BET analysis). The carbonation of

271 $Mg(OH)_2$ with a higher reactivity could have capture more CO_2 and enabled the formation of

272 nesquehonite as opposed to dypingite since nesquehonite requires a higher Mg:CO_2 molar ratio

273 as explained in the aforementioned text. This difference in the reactivity of the two samples

274 was also reflected by the absence of the residual brucite peaks in reject brine, as opposed to the

275 clearly defined uncarbonated brucite peaks observed in the $Mg(OH)_2$ slurry, as seen in Figure

276 9.
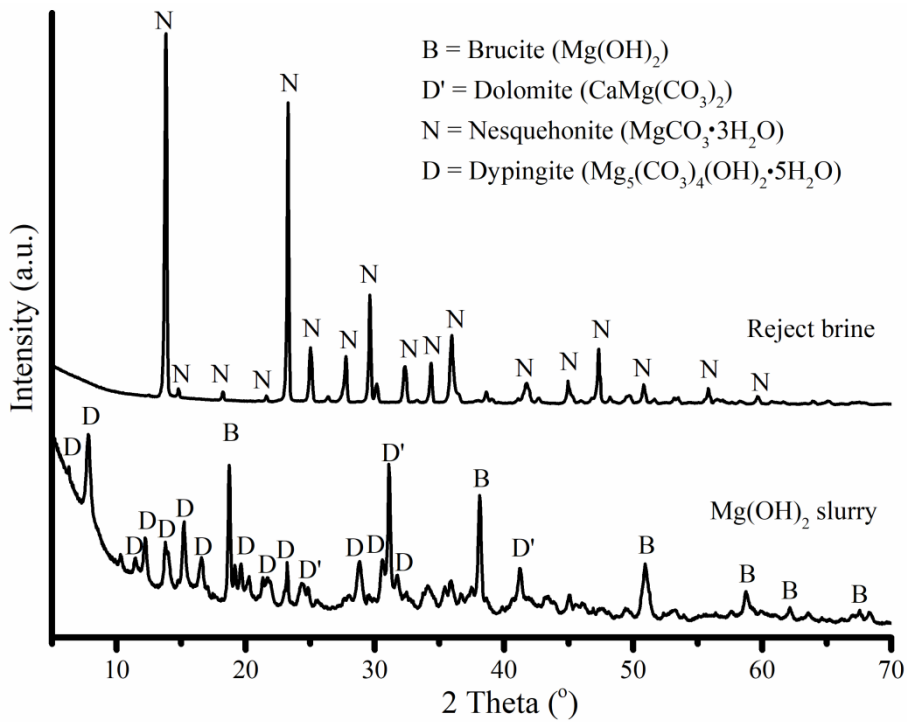
277

Figure 8 SEM images of HMCs obtained at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8, showing (a) $Mg(OH)_2$ slurry and (b) reject brine

Figure 9 XRD diffractograms of HMCs obtained from $Mg(OH)_2$ slurry and reject brine at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8

3.2.2. CO_2 capture and storage

287   The quantitative analysis of the chemical composition of HMCs and amount of CO_2 used in

288   their formation was carried out via TG/DTA and ICP-OES analyses. Figure 10 presents the

289   TG/DTA graphs of HMCs obtained from $Mg(OH)_2$ slurry and reject brine at the $Mg(OH)_2$:CO_2

290   molar ratio of 1:1 and under a pH 8. Both systems demonstrated a similar trend with three

291   stages of mass loss, which corresponded well with previous studies [36, 54, 55]. Firstly, the

292   dehydration of HMCs took place at ~100-250 °C, resulting in the loss of H_2O. The second mass

293   loss occurred between 250 and 550 °C, which was because of the decomposition of

294   uncarbonated $Mg(OH)_2$ into MgO as well as the decarbonation of HMCs, resulting in a loss of

295   H_2O and CO_2. The final mass loss observed between 550 and 700 °C was due to the

296   decomposition of dolomite and calcite, respectively, which was present as an impurity in the

297   two systems (i.e. in the form of dolomite in $Mg(OH)_2$ slurry and calcite in reject brine).

298

300   Figure 10 Typical TG/DTA curves of HMCs obtained from $Mg(OH)_2$ slurry and reject brine

301   at the $Mg(OH)_2$:CO_2 molar ratio of 1:1 under a pH of 8

302

303   The chemical composition of the final product was determined via a combination of the results

304   generated by TG/DTA and ICP-OES, where the recovery rate of Mg^{2+} was measured via ICP-

305   OES and TG/DTA was used to determine the quantity of each phase derived from XRD results.

306   Table 2 revealed the composition of the precipitate synthesized via the carbonation of $Mg(OH)_2$

307   slurry to be composed of 62.1% dypingite and 30.5% uncarbonated brucite. On the other hand,

308   the carbonation of reject brine led to a precipitate composed of 93.1% nesquehonite and 3.7%

309   uncarbonated brucite, as detailed in Table 2. These results were used in the calculation of the

310   percentage of captured CO_2, which was derived by measuring the mass of CO_2 in the HMCs

311   (i.e. dypingite/nesquehonite) divided by the initial input of CO_2 degassed to the system at the

312     $Mg(OH)_2:CO_2$ molar ratio of 1:1. The outcome of these analyses indicated that 43.7% CO_2

313   could be sequestered in the form of dypingite via the defusing CO_2 into $Mg(OH)_2$ slurry, while

314   the corresponding ratio of CO_2 sequestered in reject brine was calculated to be 82.6%. The

315   higher efficiency of CO_2 sequestration achieved via the use of reject brine was associated with

316   the increased reactivity of $Mg(OH)_2$ synthesized from reject brine. Details of the chemical

317   composition and the percentage of captured CO_2 were included in the Appendix section.

318

319   Table 2 Chemical composition of HMCs synthesized from $Mg(OH)_2$ slurry and reject brine at

320   the $Mg(OH)_2:CO_2$ molar ratio of 1:1 under a pH of 8

$Mg(OH)_2$ slurry	Mass (g)	Mass (%)	Reject brine	Mass (g)	Mass (%)
Dypingite	0.6	62.1	Nesquehonite	1.51	93.1
Brucite	0.07	30.5	Brucite	0.06	3.7
Dolomite	0.3	7.4	Calcite	0.05	3.2
CO_2 captured (%)	43.7		CO_2 captured (%)	82.6	

321

322   According to the Paris Agreement, it aims to reduce GHG emissions by 20% (*i.e.,* 7 gigatonnes

323   CO_2 emission reduction as annual anthropogenic CO_2 emissions is about 35 gigatonnes [56])

324   in order to hold the increase in the global average temperature to below 2°C above pre-

325 industrial levels [56]. According to the International Desalination Association, the global daily

326 production of desalinated water generated by 18,426 desalination plants worldwide exceeds

327 86.8 million m^3 [47]. It is estimated that an equivalent amount of reject brine is generated [47].

328 The concentration of Mg^{2+} in reject brine is around 1700 ppm, and thus around 54 million

329 tonnes of Mg could be recovered every year. With a carbon capture and storage rate of 82.6%

330 in the current study, around 45 million tonnes of CO_2 can be sequestrated annually. The

331 suggested methodology thus contributes to around 1% of the required CO_2 emission reduction

332 (*i.e.,* 7 gigatonnes) aimed in the Paris Agreement.

333

4. Summary and Conclusions

335 This study presented the influences of key parameters including the $Mg(OH)_2$:CO_2 molar ratio,

336 pH, and $Mg(OH)_2$ source on the synthesis of HMCs through the carbonation of $Mg(OH)_2$ slurry.

337 The resulting HMCs were characterized via a combination of techniques including XRD,

338 FESEM, and TG/DTA. Main conclusions drawn from this study are summarized below.

339 • The carbonation of $Mg(OH)_2$ slurry under the elevated of $Mg(OH)_2$:CO_2 molar ratio

340 resulted in the transformation of dypingite to nesquehonite.

341 • Increasing the pH from 8 to 10 was found to promote the carbonation process of

342 $Mg(OH)_2$, resulting in a higher carbonation degree.

343 • A specific "house of cards" texture, involving the formation of rosette-like dypingite

344 flakes on the surface of nesquehonite needles, was discovered under elevated pH and

345 $Mg(OH)_2$:CO_2 ratios conditions. The formation of this structure was associated with to

346 a dissolution-recrystallization-self-assembly growth mechanism as nesquehonite was

347 seen as a precursor for the further nucleation and seeding of hydromagnesite/dypingite

348 on the surface.

349 - Carbonation of $Mg(OH)_2$ slurry synthesized from reject brine led to high yield, high

350   purity, and high carbonation degree (82.6%) HMCs. Reject brine shows high potential

351   to be used for capture and long-term storage of CO_2 in the form of HMCs.

352

353 The use of compressed commercial CO_2 in the research was to provide a pure source to evaluate

354 CO_2 sequestration rate of the synthesized $Mg(OH)_2$ from desalination reject brine, which

355 served as a model cast study. In the 'real world' case, different CO_2 sources and collecting

356 methods such as CO_2 generated from factories, coal burning power plants, and municipal solid

357 waste incineration plants, may be used. However, further study is necessary to evaluate CO_2

358 sequestration efficiency of the synthesized $Mg(OH)_2$ with different CO_2 sources and collecting

359 methods. Furthermore, it is necessary to evaluation of the mass, energy, reagents, wastes that

360 come into play in the global process from the life cycle and life cycle cost viewpoint of HMCs

361 synthesized from reject brine.

362

363 **Acknowledgement**

364

368

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521

522

Appendix – Chemical composition of HMCs and CO$_2$ captured percentage

<u>HMCs synthesized from Mg(OH)$_2$ slurry</u>

Initially, 0.82 g Mg(OH)$_2$ is added into 200 ml distil water at a Mg(OH)$_2$/CO$_2$ molar ratio of 1. Assuming a purity of 92% for Mg(OH)$_2$, the remaining 8% impurity is dolomite which does not react or dissolve in the solution. The final precipitates consist of uncarbonated brucite, dypingite and dolomite as supported by XRD results (Figures 5 and 7), which after calcination decompose into MgO, MgO, and CaO·MgO, respectively. Let x and y denote the weights of the uncarbonated brucite and dypingite, respectively. The weight of dolomite is 0.066 g calculated based on 8% of the initial sample weight (i.e., 0.82 g). Residues after the TG/DTA test is 50.6% (Figure 10). Based on the given information, the following equation can be established.

$$\frac{\frac{40}{58}x + \frac{40}{96.8}y + 0.82 \times 0.08 \times \frac{96}{184}}{x + y + 0.82 \times 0.08} = 0.506 \tag{A1}$$

Furthermore, the concentration of Mg^{2+} in the residue was measured to be 308.4 ppm (0.308 g/L) as shown in Table 3. Thus,

$$\frac{\frac{24}{58}x + \frac{24}{96.8}y}{0.2} = \frac{\frac{24}{58} \times 0.82 \times 0.92}{0.2} - 0.308$$
$$\text{(A2)}$$

By solving the two Eqns. (A1) and (A2), the weights of uncarbonated brucite and dypingite are calculated to be 0.27 g and 0.55 g, respectively. The sum of the weights of uncarbonated brucite, dolomite and dypingite was calculated to be 0.89 g, which was higher than 0.46 g of the weighted precipitate as shown in Table 3. This was mainly due to the weight losses during the process of separating the samples from the solution and grinding. The captured CO$_2$ percentage was calculated by measuring the weight of CO$_2$ in the HMCs (dypingite) divided by the initial

547 input of CO_2 degassed to the system at the $Mg(OH)_2/CO_2$ molar ratio of 1, which was calculated

548 to be 43.7% in consideration of 8% impurity.

549

550 Table A1 Concentration of Mg^{2+} in the residue and final weight of HMCs obtained from

551 $Mg(OH)_2$ slurry and reject brine

	Mg^{2+} in the residue solution (ppm)	Weight of solids (g)
$Mg(OH)_2$ slurry	309.9 ± 5.2	0.46 ± 0.20
Reject brine	161.5 ± 9.1	1.24 ± 0.14

552

553 <u>HMCs synthesized from reject brine</u>

554 The same principal applies to HMCs synthesized from the reject brine. The $Mg(OH)_2$ sample

555 precipitated from reject brine in the first step contains 6.3% calcite as the impurity which has

556 been detailed in [50]. The XRD result confirms that HMCs after carbonation of $Mg(OH)_2$

557 consisted of nesquehonite, uncarboned brucite and calcite, which after calcination decompose

558 into **MgO, MgO and CaO**, respectively. Let x and y denoted the weights of uncarboned brucite

559 and nesquehonite, respectively. The weight of calcite is 0.052 g calculated based on 6.3% of

560 the initial sample weight (i.e., 0.82 g). Residues after the TG/DTA test is 31.3% (Figure 10).

561 Based on the given information, the following equation can be established.

562
$$\frac{\frac{40}{58}x + \frac{40}{138}y + 0.82 \times 0.063 \times \frac{56}{100}}{x + y + 0.82 \times 0.063} = 0.313 \tag{A3}$$

563 Furthermore, the concentration of Mg^{2+} in the residue brine was measured to be 159.3 ppm

564 (0.159 g/L) as shown in Table 3. Thus,

565
$$\frac{\frac{24}{58}x + \frac{24}{138}y}{0.2} = \frac{\frac{24}{58} \times 0.82 \times 0.937}{0.2} - 0.159 \tag{A4}$$

566

567 By solving the two Eqns. (A1) and (A2), the weights of uncarbonated brucite and nesquehonite

568 are calculated to be 0.06 g and 1.51 g, respectively. The captured CO_2 percentage was

569 calculated by measuring the weight of CO_2 in the HMCs (nesquehonite) divided by the initial

570   input of CO_2 degassed to the system at the $Mg(OH)_2/CO_2$ molar ratio of 1, which was calculated

571   to be 82.6% in consideration of 6.3% impurity. A successful sequestration of CO_2 into reject

572   brine as HMCs was therefore achieved to obtain an efficiency as high as 82.6%, which was

573   significantly improved compared to the $Mg(OH)_2$ slurry.

574