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1 Editorial

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3 A new focus on legacy pollutants: Chlorinated Paraffins (CPs) and  
4 Polychlorinated Naphthalenes (PCNs)

5

6 With volume production spanning almost a century, chlorinated paraffins (CPs) are industrial  
7 chemicals that are of increasing concern to the environment and to human health.  
8 Polychlorinated naphthalenes (PCNs) are legacy contaminants that occur in the environment,  
9 food chains and human tissues and contribute to the burden of dioxin-like effects such as  
10 carcinogenicity, hepatotoxicity, teratogenicity, embryotoxicity, etc. In recognition of these  
11 concerns, and following structured risk assessments, short-chain CPs (SCCPs, C<sub>10</sub>-C<sub>13</sub>) and  
12 PCNs have recently been listed for elimination, in Annex A of the Stockholm Convention.  
13 This followed an earlier listing for SCCPs as a priority hazardous substance for control under  
14 the European Union (EU) Water Framework Directive. The importance of these  
15 developments has not been lost on scientists who have researched an increased volume of new  
16 information in these fields. The Dioxin 2018 symposium in Krakow devoted a full day to the  
17 dissemination of the latest findings, with 18 oral presentations and 8 posters covering  
18 analytical aspects, occurrence in the environment, materials and food, and the toxicological  
19 effects of these pollutants.

20 Chemically, CP products are complex isomeric mixtures of several thousands of individual  
21 compounds having carbon chain lengths ranging from C<sub>10</sub> to C<sub>28</sub>. Reliable analytical  
22 determination is one of the most intractable barriers to the accurate measurement of CP  
23 occurrence, a recognised drawback that was illustrated in the inter-laboratory studies by  
24 Krätschmer and Schächtele (*Chemosphere* 234, 252-259). The issue is compounded by the  
25 ambiguity in defining CP analytes – a necessary first step in reliable analytical determination.

26 The Stockholm convention lists SCCPs in its annex, but this relates to commercial products.  
27 An innumerable selection of different “SCCP” products of varying compositions were  
28 manufactured globally, and the environmental legacy of widespread historical usage that is  
29 observed in environmental media is a complex integral of these mixtures. This integral is  
30 further modified by transformative processes such as modification during usage, selective  
31 rates of evaporation, photochemical and microbial degradation (Heeb et al. *Chemosphere* 226,  
32 744-754), etc. Reported occurrences in food of animal origin (Krätschmer et al., *Chemosphere*  
33 227, 630-637; Labadie et al., *Chemosphere* 223, 232-239; Jiang et al., *Chemosphere* 229, 358-  
34 365) imply additional transformation through metabolic processes which further enhance the  
35 complexity of the observed profiles. It is therefore unsurprising that the profiles for SCCPs  
36 (and other CPs) observed during analysis do not correspond to individual commercial  
37 products, and perhaps more relevantly, to analytical standards.

38 In general, early insights into the toxicological effects of CPs were based on the use of  
39 standard mixtures that reflected the commercial products. As CP residues in real foods and  
40 animal tissues have never been completely characterised (but are clearly modified integrals of  
41 different mixtures), it would prove difficult for a human exposure based risk assessment to  
42 correlate occurrences to the reported effects. Another pressing issue is that of current CP  
43 manufacture and use. The most recent literature suggests that shining a regulatory spotlight on  
44 SCCPs, has resulted in a shift to the use of medium- (MCCPs, C<sub>14</sub>-C<sub>17</sub>) and long-chain CP  
45 (LCCPs, > C<sub>18</sub>) mixtures. Despite some investigations on MCCPs within the EU and North  
46 America there is still a lower level of knowledge on the toxicity of these products, particularly  
47 the LCCPs. In the long term, regulation of SCCPs on their own is unlikely to address potential  
48 risks arising from these other mixtures and their breakdown products. Further, CP mixtures  
49 are known to contain other chlorinated contaminants as by-products, such as polychlorinated  
50 biphenyls (PCBs), PCNs and chlorinated dioxins and furans (PCDD/Fs), and may also give

51 rise to these as by-products during combustion (Matsukami and Kajiwara *Chemosphere* 230,  
52 164-172). Clarification is required as to whether the toxicological effects that have been  
53 reported were directly attributable to CPs, and were not influenced, at least in part, by the  
54 presence in the test mixtures of such by-products which show more sensitive toxicological  
55 endpoints than CPs.

56 As the assessment of exposure is an integral component of human health risk assessment, a  
57 useful first step would be to characterise the CP profiles observed in foods as these are  
58 expected to constitute an important exposure pathway (as observed with other similar  
59 halogenated contaminants). This should include food packaging materials which have also  
60 been shown to contain CPs (Wang et al., *Chemosphere* 225, 557-564). Recent advances in  
61 instrumentation that allow qualitative homologue group characterisation of CP occurrence  
62 may prove a useful tool in the identification and mapping of groups that predominate in  
63 “typical” profiles for different food types. As many household and workplace materials are  
64 known to contain CPs, a similar approach to characterising these occurrences would yield  
65 information on other possible exposure pathways.

66 This characterisation would have two immediate advantages – it would provide a more  
67 focussed approach to toxicological studies by allowing the targeting of relevant (occurring)  
68 homologue groups and help identify groups that elicited more potent responses, and also  
69 provide direction to the analytical effort by indicating a qualitative definition of the analytes.  
70 The characterisation would also allow the formulation of more relevant standard CP mixtures  
71 that correlate to a greater extent with observed profiles, thus aiding quantitative  
72 determination. In this context specific single chain length mixtures are currently being  
73 synthesised and characterised (Sprenkel et al., *Chemosphere* 228, 762-768). The discrepancies  
74 observed in the most recent inter-laboratory comparisons (Krätschmer and Schächtele,  
75 *Chemosphere* 234, 252-259) underline the requirement for representative standards, but also

76 highlight the need for a robust and harmonised approach to the quantitation procedures  
77 applied to the identified CP homologue groups. Until there is progress on these issues, the  
78 expression of CP concentrations as total CP (either combined, or if analytical advances allow,  
79 speciated into short, medium and long chain) would be a sensible interim measure, allowing  
80 the generation of much needed occurrence data and laying the groundwork for future control  
81 and regulation efforts.

82 For PCNs, information from emerging research continues to define the issues surrounding  
83 these contaminants. An increasing amount of recent literature that speciates PCN occurrences  
84 in environmental media and foods, by individual congeners, provides further information on  
85 the persistence and fate of these chemicals, decades after production ceased. The historical  
86 and continuing human exposure arising from PCN occurrence in foods and dietary  
87 supplements (Falandysz et al., *Chemosphere* 231, 240-248; Zhihua et al., *Chemosphere* 230,  
88 559-566) underlines the persistence of PCN congeners and the enduring legacy of this  
89 contamination in marine regions from where current fish supplies continue to be sourced.  
90 New insights into the environmental behaviour and chemistry of individual congeners help to  
91 explain observed patterns in environmental media and the resulting occurrence, particularly in  
92 marine products.

93 Relative to CPs, the analytical determination of PCNs is at an advanced level with reliable  
94 measurement of individual congeners allowing behavioural studies of selected compounds. It  
95 is particularly encouraging to see new work that adds to the body of toxicological insights into  
96 PCN disposition in animal tissues and the effects on reproductive processes (Kilanowicz et  
97 al., *Chemosphere* 226, 75-84; Kilanowicz et al., *Chemosphere* 228, 577-585). The dioxin-like  
98 behaviour of some PCN congeners has been recognised for several years, but the  
99 identification of other toxicological effects such as disruption to haemostasis parameters such

100 as clot formation and fibrinolysis, adds to the growing evidence of requirement for future  
101 regulatory action.

102 Although the majority of the work presented at Dioxin 2018 included targeted studies with  
103 specific outcomes directed to ultimately investigating the environmental, human exposure and  
104 health effects of PCNs and CPs, it is important that this collated dissemination is viewed  
105 within a wider context. Both of these contaminant classes are mass produced anthropogenic  
106 products that have seen, often unrestricted, usage for the best part of a century. However, the  
107 volume of pertinent literature is relatively small in comparison to other similar contaminants  
108 such as PCBs, PCDDs/Fs and flame retardants. In the case of CPs, the combination of a lack  
109 of widespread recognition combined with the real difficulty with analytical access is a clear  
110 factor. For PCNs, the similarity of chemical behavior and effects to the more widely  
111 produced PCBs has overshadowed the potent toxicological response of these chemicals.  
112 However, in the light of the current re-evaluation of PCB toxicity, particularly PCB 126, the  
113 contribution of PCNs to the cumulative dioxin-like toxicity could potentially become more  
114 significant, engendering more interest in the regulation of these contaminants as well.

115 The inclusion of both these classes of contaminants within the Stockholm convention listing  
116 has been followed by regional interest, e.g. within the EU, which has set up specific working  
117 groups to address human exposure through the occurrence of these chemicals in food. Both of  
118 these measures, provide direction to the task for scientists to facilitate and generate  
119 information that will ensure that the remaining challenges and risks to human health are  
120 characterized and are available for policy making.

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122 contributing authors, and especially to the reviewers who collectively ensured a high standard  
123 of scientific dissemination.

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