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# Report of the Working Group on the Celtic Seas Ecoregion (WGCSE) 

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## International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer

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## 1 General

The report has no Executive Summary.

A discussion of historical data available is discussed in Section 2.1 to 2.5 of WGNSDS 2008 and Section 1.3 of WGSSDS 2008 (ICES, 2008a,b). There have been no substantive changes to available data or work up methodologies this year. The methods employed by the WG are described in each stock annex and Sections 2.6 to 2.11 of WGNSDS 2008 and Section 1.4 of WGSSDS 2008 (ICES, 2008a,b).

Biological sampling levels by country and stock are summarised in Table 2.1. The sampling levels for 2008 are, in general, similar to those in 2007. Deficiencies in sampling (if any) are discussed in the relevant stock section.

## References

ICES. 2010. Report of the Working Group on the Assessment of Northern Shelf Demersal Stock (WGNSDS), 15-21 May 2008, Copenhagen, Denmark. ICES CM 2008/ACOM:08. 756 pp.

ICES. 2010. Report of the Working Group on the Assessment of Southern Shelf Demersal Stocks (WGSSDS), 26 June-5 July 2007, ICES Headquarters, Copenhagen. ICES CM 2007/ACFM:28. 675 pp .

### 2.1 MSY estimation for fin-fish stocks

The general approach of WGCSE is outlined as follows: WGCSE used ADMB to explore the S-R, fishery selection, and growth potential data for fin-fish stock where assessment data were available. Based on an analysis of the uncertainty of the estimated parameters, the AICc value and the coefficient of variation the fits and estimates from various S-R models (Beverton and Holt, Ricker and smooth hockey stick) the most plausible S-R relationships were used for the estimation of Fmsy. In many cases the Fmsy estimates were equally well determined by each model but often differing S-R models result in differed in the absolute values.

Where this was the case WGCSE concluded that as result of the equivalence in the precision of the estimates determined from each model fit, no definitive value of Fmsy can be defined. In such cases the range of estimated fishing mortalities should be used as the basis for the management advice for the stock. For example fishing mortalities in the range 0.19-0.36 are consistent with maximising long-term yield for plaice in VIIfg, no value in the range is considered more appropriate than any other. The advice could be framed using the maximum of this range as an upper bound for $\mathrm{F}_{\mathrm{msy}}$. In the example of plaice VIIfg the ICES MSY framework implies that fishing mortality be reduced to 0.36 or lower.

### 2.2 MSY estimation for Nephrops stocks

The different Nephrops stocks (Functional Units, FUs) for which ICES delivers advice cover a wide range of fisheries including single, twin, triple and even quadruple trawls, creeling (potting), with activity covering inshore and offshore grounds. The timing of these fisheries varies, which due to the different emergence patterns of the different sexes due to moulting and egg-brooding, leads to very different relative exploitation rates (between the sexes) in different FUs. Local ecosystem type is also highly variable with a range of Nephrops densities, different composition and density of organisms competing for space as well as different assemblages of predators. Ground types also cover a wide range including large contiguous sediment beds, fragmented patches of suitable sediment in rocky areas, shallow sea-lochs and
patches of mud on relatively deep shelf-edges. Given these differences in fishery and ecology it is inevitable that estimates of the exploitation rate leading to long-term MSY will vary between the FUs, the difficulty for scientists is how to estimate these rates given the inherent difficulty in assessing crustacean stocks, for which no practical method routine of age determination is available. Some assessments take the observed length frequency data and slice it into age-classes according to the von Bertalanffy growth parameters. These numbers-at-age are then taken forward into standard stock-assessment packages. This practice was ceased in 2005 within this Group due to concerns over both the reliability of reported landings in some FUs (particularly the UK fisheries) and the use of the 'pseudo' age-structured data in an age-based assessment. As a result of this, no dynamic population model is fitted to the data and consequently there are no estimates of spawning stock and recruitment which are fundamental to the determination of $\mathrm{F}_{\text {msy }}$ and proxies for $\mathrm{F}_{\text {msy }}$ must therefore be sought. WKFrame (ICES 2010) made several recommendations for defining $\mathrm{F}_{\text {msy }}$ proxies where no direct estimation of $\mathrm{F}_{\text {msy }}$ was possible (i.e. for stocks for which there is no analytic assessment, but length- or age-structured catch data are available). The suggested approach focused on per-recruit analysis with the following guidelines:

- Use input parameters which reflects the current situation (selection and discard ogive, maturity and weight-at-age/length)
- If there is clear peak at low F in the YPR analysis and no evidence of recruitment dependence on biomass, then $F_{\text {max }}$ may be an appropriate proxy.
- Where $\mathrm{F}_{\max }$, is undefined then $\mathrm{F}_{0.1}$ might be considered as a 'lower bound' to the range of F suitable for $\mathrm{F}_{\mathrm{msy}}$, as it is assumed to be low risk.
- Spawning biomass per recruit analysis should be routinely evaluated in addition to YPR. There is not a single level of \% SPR that is optimal for all stocks and the proposal for $\mathrm{F}_{\text {msy }}$ should include some consideration of life history. Further studies by Clark (1991; 1993) concluded that F35\% and higher were robust proxies for $\mathrm{F}_{\mathrm{msy}}$, considering uncertainty in stockrecruitment functions and or recruitment variability.
- Conduct a sensitivity analysis to the input parameters and consider the variability of estimates over time.

Within the Celtic sea areas, assessment of Nephrops stocks falls into three categories, those with TV surveys, those monitored by lpue/mean size and those with only landing information. Only for those stocks with TV surveys is the catch advice determined by an exploitation rate, advice for the other stocks is based on changes to landings. For those stocks with a TV survey, the Harvest Rates (removals divided by abundance as estimated by the TV survey) associated with fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ were estimated at the 2009 benchmark meeting WKNeph (ICES 2009). In response to the recommendations of WKFrame, estimates of $\mathrm{F}_{35 \% \mathrm{SpR}}$ and the corresponding Harvest Rate have also been determined and these estimates typically lie between the estimates of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max }}$. Suggestions for a TV-abundance based proxy for $\mathrm{B}_{\text {trigger }}$ have been made on the basis of the lowest observed TV-abundance (median survey value) unless the stock has shown signs of stress at a higher TV-abundance in which case this value becomes $B_{\text {trigger }}$.

The remaining challenge is determining which $\mathrm{F}_{\text {msy }}$ proxy is appropriate for which stock and this becomes an exercise in expert judgment based upon knowledge of the fishery and the ecosystem. The implications for exploitation rate can vary considerably depending upon which proxy is chosen ( $\mathrm{F}_{0.1}, \mathrm{~F}_{35 \% \mathrm{SpR}}$ or $\mathrm{F}_{\max }$ ) and whether to ac-
count for the differences in relative exploitation rate between the sexes. Given that there is often a distinct difference in the exploitation rate between the two sexes (males>females) it is usually impossible to simultaneously achieve the target fishing mortality on both sexes (i.e. the stock cannot be fished such that both the male and female YPRs are maximised simultaneously). The following text table shows the Fmultipliers required to achieve various $\mathrm{F}_{\mathrm{msy}}$ proxies for the sexes of a typical Nephrops stock (FU 8 in this example), the Harvest Rates which correspond to those F multipliers and the resulting level of spawner-per-recruit expressed as a percentage of the virgin level.

|  |  | Fbar(20-40 mm) |  |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fmult | Male | Female |  | Male | Female | Combined |
| $\mathrm{F}_{0.1}$ | Male | 0.2 | 0.13 | 0.06 | 7.47 | 42.33 | 64.50 | 51.72 |
|  | Female | 0.43 | 0.29 | 0.13 | 14.23 | 22.96 | 44.80 | 32.21 |
|  | Combined | 0.24 | 0.16 | 0.07 | 8.75 | 37.29 | 60.04 | 46.92 |
| $\mathrm{F}_{\text {max }}$ | Male | 0.36 | 0.24 | 0.11 | 12.31 | 26.94 | 49.50 | 36.49 |
|  | Female | 0.81 | 0.54 | 0.24 | 23.38 | 12.11 | 28.95 | 19.24 |
|  | Combined | 0.46 | 0.31 | 0.14 | 15.03 | 21.55 | 43.02 | 30.64 |
| F35\%SpR | Male | 0.27 | 0.18 | 0.08 | 9.67 | 34.13 | 57.04 | 43.83 |
|  | Female | 0.63 | 0.42 | 0.19 | 19.28 | 15.79 | 34.96 | 23.91 |
|  | Combined | 0.39 | 0.26 | 0.12 | 13.15 | 25.10 | 47.38 | 34.53 |

The yield-per-recruit and spawner-per-recriut plots for this stock are shown in Figure 2 , emphasizing the disparity in f-multipliers required to achieve $\mathrm{F}_{\text {max. }}$. The general tradition in fisheries science is to concentrate on the mortality on females because in a freely distributing population, one male should be able to fertilise several females and therefore a higher exploitation rate on males should not affect spawning potential. Nephrops are slightly different in that the adults have a fairly limited range of movement (100's of metres) and therefore very low densities of males could result in sperm limitation. Ensuring that the fishing mortality target on males is not exceeded will usually result in an under-utilisation of the females, but due to the faster growth rate of males the under-utilisation of total yield is not likely to be large. The alternative, of trying to achieve $\mathrm{F}_{\mathrm{msy}}$ on females, carries a potentially serious risk to the production of future recruits and may result in very high exploitation of males. A the use of a combined $\mathrm{F}_{\mathrm{msy}}$ (or proxy thereof) would obviously deliver higher long-term yield than either of the two separate sex values but the implication for male stock level should be noted. The Working Group suggested that a combined sex Fmsy proxy should be considered appropriate provided that the resulting percentage of virgin spawner-per-recruit for males does not fall below $20 \%$. In such a case the male $\mathrm{F}_{\text {msy }}$ proxy should be picked over the combined proxy.

In cases where recruitment rates are typically low and/or highly variable then a more cautious $\mathrm{F}_{\text {msy }}$ proxy would be appropriate as the stock may have reduced resilience to periods of poor recruitment and in this case $\mathrm{F}_{0.1}$ is recommended. Conversely where recruitment rates are considered to be regularly high and the stock appears to have supported a harvest rate at or above $\mathrm{F}_{\text {max, }}$ (or in the case of a short TV time-series a particular landing level) without showing signs of recruitment overfishing, then $\mathrm{F}_{\max }$ is recommended. In all other cases $\mathrm{F} 35 \% \mathrm{SpR}$ should deliver high long-term yield with a low probability of recruitment overfishing and is recommended as the "default" value.

In order to assist communication of the decision process the following bullet list is suggested as a standard checklist for describing the rationale behind the choice of a particular $\mathrm{F}_{\text {msy }}$.

- Describe the absolute density. Is it high (i.e. $>1$ per $m^{2}$ ), medium (i.e. 1.0-0.2 per $m^{2}$ ) or low (i.e. $<0.2$ per $m^{2}$ )
- Variability in density. Is there large interannual variability, spatial complexity?
- Understanding of biological parameters. Is the growth rate particularly fast or slow, high or low estimates of natural mortality?
- Fishery timing and operation. Is there a strong seasonal pattern leading to different exploitation rates on the sexes, does this pattern vary much between years?
- Observed Harvest Rate or landings compared to stock status. Is the harvest rate consistently around or above $F_{\text {max }}$ ? Have landings been stable? Have the indicators of stock status shown signs of difficulty?

Accompanying this text should be a table listing the $\mathrm{F}_{\text {msy }}$ proxies $\mathrm{F}_{\text {max }}, \mathrm{F}_{35 \%} \mathrm{Fspr}_{\mathrm{p}}$ and $\mathrm{F}_{0.1}$ for males and females, the Harvest Rates they correspond to along with the implied \%spawner-per-recruit for males and females.

Following changes to UK legislation in 2006 the reliability of UK landings data is considered to have significantly improved (representing $\sim 80 \%$ of the landings). Provided that this is both true and continues into the future, assessment scientists will eventually have data which could be used to parameterise dynamic stock assessment models which in turn will enable estimation of $\mathrm{F}_{\text {msy }}$ directly rather than have to rely upon proxies thereof. Until this point the decision of which $\mathrm{F}_{\text {msy }}$ proxy is suitable for which FU will inherently be a subjective process but the process outlined above should provide sufficient justification to support the decision.



Figure 2.1. Yield-per-recruit and spawning-stock biomass-per recruit for males, females (dotted line) and combined (bold) with $F_{\max }$ and $\mathrm{F}_{35 \%}$ spr reference points.

Table 2.1. Biological sampling levels by stock and country.


### 3.1 West of Scotland overview

There is no overview.

### 3.2 Cod in Subarea Vla

Cod in Division VIa is included in the EU long-term management plan for cod stocks and the fisheries exploiting those stocks (Council Regulation (EC) 1342/2008). An update assessment was conducted this year by the WG.

ICES advice applicable to 2009

## Exploitation boundaries in relation to existing management plans

The management plan is not explicit about the level of reduction in the catch when the stock is below Blim. Furthermore, due to the uncertainty in the level of fishing mortality, ICES is not in a position to give quantitative forecasts. Simulations conducted in 2006 showed that fishing should be closed for 3 years in order to bring SSB above Blim.

## Exploitation boundaries in relation to precautionary considerations

Given the low SSB and low recruitments in recent years, it is not possible to identify any non-zero catch which would be compatible with the precautionary approach.

## ICES advice applicable to 2010

## Single stock exploitation boundaries

ICES evaluated the long-term management plan and has not yet been able to confirm that it is precautionary. Considering the options below, ICES advises on the basis of exploitation boundaries in relation to precautionary considerations that no fishing should take place on cod in Division VIa.

## Exploitation boundaries in relation to existing management plans

Due to the uncertainty in the level of fishing mortality, ICES is not in a position to give quantitative forecasts. Given the stock status it is likely that the stock will fall into the category defined in Article 9.a of the plan which implies a $25 \%$ TAC reduction.

## Exploitation boundaries in relation to precautionary considerations

Given the low SSB and low recruitments in recent years, it is not possible to identify any non-zero catch which would be compatible with the precautionary approach.

### 3.2.1 General

Stock definition and the management unit
General information about the stock can be found in the Stock Annex and an overview of the fisheries West of Scotland can be found in Section 3.1. The assessment unit is VIa and a TAC is set for ICES Areas VIa and Vb (EC waters). The 2009 and 2010 TACs for cod in the management unit were 240 t and 240 t respectively.

## Management applicable to 2009 and 2010

The minimum landing size of cod in the human consumption fishery in this area is 35 cm . Before 2009 a TAC was set for ICES Subarea VI and EC and international waters of ICES Subareas XII and XIV and Subdivision Vb1. From 2009 a TAC for VIa and Vb 1 was given.

TAC for 2009

| Species: Cod <br> Gadus mortua |  | Zone: | VE EC waters of Vb ; EC and international waters of XII and XIV <br> (COD/561214) |
| :---: | :---: | :---: | :---: |
| Belgium | 0 |  |  |
| Germany | 4 |  |  |
| France | 48 |  |  |
| Ireland | 68 |  |  |
| United Kingdom | 182 |  |  |
| EC | 302 |  |  |
| TAC | 302 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

Special conditions
Within the limits of the abovementioned quotas, no more than the quantities given below may be taken in the ICES zones specified

|  | Vla; EC waters of Vb <br> $\left(\mathrm{COD}^{*} 5 \mathrm{SBC6A}\right)$ |
| :--- | :---: |
| Belgium | 0 |
| Germany | 4 |
| France | 38 |
| Ireland | 54 |
| United Kingdom | 144 |
| EC | 240 |

TAC for 2010

| Species: Cod <br> Gadus morhua |  | Zone: | $\mathrm{Vla} ; \mathrm{EU}$ and international waters of Vb east of $12^{\circ} 00$ <br> w <br> (COD/5B6A-C) |
| :---: | :---: | :---: | :---: |
| Belgium | 0 |  |  |
| Germany | 4 |  |  |
| France | 38 |  |  |
| Ireland | 53 |  |  |
| United Kingdom | 145 |  |  |
| EU | 240 |  |  |
| TAC | 240 |  | Analytical TAC |

Technical measures applicable to the West of Scotland, including those associated with the cod recovery plan in force in 2008 (Council Regulation No. 423/2004), the cod long-term management plan in force from 2009 (Council Regulation No. 1342/2008) and the Restrictions on fishing for cod, haddock and whiting in ICES Zone VI contained in Council Regulation No. 43/2009 (Annex III paragraph 6), are described in Section 3.1.

## The fishery in 2009

Cod is believed to be no longer targeted in any fisheries now operating in ICES Division VIa. The table of official landings statistics is given in Table 3.2.1. This indicates the full TAC was not taken, however, the catch information provided to ICES for the assessment give landings close to the TAC (Table 3.2.2).

Because of restrictive TACs, seasonal/spatial closures of the fishery, and effort restrictions based on bycatch composition the probability of misreporting and underreporting of cod in the past is considered to have been high. From 2006 the Registration of Buyers and Sellers legislation in the UK and Sales Notes management system in Ireland are considered to have reduced to low levels underreporting (see Section 3.1). Area misreporting, however, is still believed to take place in the UK. Area misreporting will, for example, see cod caught in VIa declared as taken from the Faroe region or ICES Area IVa. The UK and Irish legislation introduced in 2006 is also believed responsible for a significant increase in discards starting in 2006. Since 2006, the estimated weight of discards has exceeded landings (Table 3.2.2), and discarding has taken place over an increased range of age groups (Tables 3.2.6 and 3.2.7 and Figure 3.2.1). Discard numbers as a percentage of catch numbers-at-age for 2009 are shown in the following text table:

| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 +}$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \% catch <br> discarded | 99.8 | 95.7 | 94.8 | 82.1 | 0.0 | 88.0 | 0.0 |

The absolute level of numbers discarded from the 2005 year class at age 1 in 2006 through to age 4 in 2009 have been high relative to the same age class from adjacent cohorts (Table 3.2.6). Estimates of catches (landings plus discards) derived from observer programme and logbook data are almost seven times higher than reported landings.

Tables and figures of total effort by the fleets operating in Division VIa can be found in Section 3.1.

### 3.2.2 Data

An overview of the data provided and used by the WG is provided in the following text table:

|  | Commercial Data |  |  |  | Survey Data |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings |  | Discards |  | cpue at age |  |  |  |
|  | No. at age | Wght. <br> at age | No. at age | Wght. <br> at age | $\begin{aligned} & \text { ScoGFS- } \\ & 1 Q \end{aligned}$ | ScoGFS- <br> 4Q | IreGFS | IGFS |
| Available | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1985- \\ & 2010 \end{aligned}$ | $\begin{aligned} & 1996- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1993- \\ & 2002 \end{aligned}$ | $\begin{aligned} & 2003- \\ & 2009 \end{aligned}$ |
|  | $\begin{aligned} & \text { Ages : } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages : } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 1-7 \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 0-8 \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 0-3 \end{aligned}$ | $\begin{aligned} & \text { Ages : } \\ & 0-3 \end{aligned}$ |
| Used | $\begin{aligned} & 1978- \\ & 1994 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 1994 \end{aligned}$ | $\begin{aligned} & 1978- \\ & 2009 \end{aligned}$ | $\begin{aligned} & 1985- \\ & 2010 \end{aligned}$ | NOT <br> USED | NOT <br> USED | NOT <br> USED |
|  | $\begin{aligned} & \text { Ages : } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages : } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages : } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 1-7+ \end{aligned}$ | $\begin{aligned} & \text { Ages: } \\ & 1-6 \end{aligned}$ |  |  |  |

A plot of log catch curve gradient derived from commercial catch data (landings plus discards) is shown in Figure 3.2.2. The trend in gradients over time appear fairly consistent between the age ranges considered ( $2-5,2-4$ and $3-5$ ) except for the most recent two cohorts (2004 and 2005 cohorts). The implication from the figure is of an increasing rate of mortality for cohorts spawned during the 1990s, a considerable reduction in mortality for the 2002, 2003 and 2004 cohorts, but a return to a higher mortality rate for the 2005 cohort. Landings and discard data numbers-at-age are, however, only included in the assessment up to 1994 because of concerns over deteriorating quality of landings data.

Annual mean weights-at-age in landings, discards and catch are given in Tables 3.2.5, 3.2.7 and 3.2.9. Weights-at-age for the stock are still required to obtain biomass estimates and so the full series of stock weights are used. Figure 3.2.1 shows the mean weights-at-age in the landings and discards. There is no evidence of a trend in weight-at-age for ages 1 and 2 for VIa cod landings, but some evidence of a gradual long-term decline at age 3 and above. Mean weight-at-age of discarded fish at age 2 has increased in recent years.

Raised discard numbers-at-age are given in Table 3.2.6. Discards for the international fleet were raised from Scottish observations (see also the Stock Annex) Observer coverage in 2009 (number of trips) is detailed in the following text table:

|  |  | AREA VI |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Other trawlers | Nephrops trawlers | Total |
| 2008 | 9 | 8 | 17 |
| 2009 | 10 | 22 | 32 |

Increased discards from 2006 are considered an indicator of the combined effect of restrictive quotas and new regulation. The larger 2005 cohort can be tracked through the discards. A consequence of the current assessment model configuration-discard proportions modelled for ages one and two only and discard information not used after 1994-is that the change in discarding practices from 2006 as shown in Tables 3.2.2 and 3.2.6 have no influence on the final assessment.

All available survey data are given in Table 3.2.3, with the data used in the assessment highlighted in bold. Survey descriptions are given in the stock annex. Figure 3.2.3 shows cpue by survey haul from 2009 for the ScoGFS-4Q and IGFS surveys and from 2010 for the ScoGFS-1Q survey. The data from the Scottish surveys show cpue or ages $1+$, that from the Irish survey a proxy for fish at ages $1+$ (fish at lengths $>23$ $\mathrm{cm})$. The quarter four surveys show catches of cod in the northern part of the region (north Minches and north of 58.5 degrees N ) and in the southern part of the region (off the north coast of Ireland and along the shelf edge south of 56 degree N ) but mostly zero returns in the intervening latitudes. This pattern has been relatively consistent over the years 2007-2009. Since 2000 the ScoQ1 survey has caught very few cod in the southern region especially west of 7 degrees west (see also Figure A9.3 in the Stock Annex).

Figure 3.2.4 shows log catch curves for the ScoGFS-1Q survey. It shows a strong "hook" at the younger ages, with abundance at age two often higher than at age one. The index of the 2005 year class has, however; also increased from age 2 to age 3 and the survey's ability to track recent cohorts is poor due to the low abundance and catch rates.

Values for natural mortality ( 0.2 for all ages and years) and the proportion of fish ma-ture-at-age are unchanged from the last meeting. The proportion of F and M acting before spawning is set to zero.

A study by the sea mammal research unit (SMRU) on seal predation has indicated that seal predation on cod probably constitutes significant natural mortality over and above values assumed in the assessment. A working document looking at the significance of seal predation to perceptions of the VIa cod stock was submitted to WGNSDS_08 and work is ongoing for incorporation into the VIa cod benchmark. Any increase in predation mortality would be incorporated into estimates of total mortality by the stock assessment model.

A plot of $\log$ catch curve gradient derived from the ScoGFS-1Q data is shown in Figure 3.2.5. For cohorts after 1995 index values of zero have sometimes been recorded at age five. For the age ranges considered ( $2-5,2-4$ and $3-5$ ) this means the slope has not always been fitted to data from all the ages indicated. There is little consistancy in results between age ranges chosen and this appears to worsen after the 1995 or 1996 cohort. The series for ages $2-5$ seems more stable than the others in this later period. In contrast to the commercial data the result for the 2005 cohort shows a large decline in mortality rate on this cohort. Overall, information on mortality trends from all survey series (including the ScoGFS-1Q) appears weak wich results in wide confidence bands around the mortality signal.

### 3.2.3 Historical stock development

This update assessment uses a TSA run as outlined in the Stock Annex.
Model settings and input parameter settings for the final run are given in Table 3.2.10 and final parameter estimates from the TSA run are given in Table 3.2.11, alongside final run estimates for VIa cod from previous WGs. Standardised prediction errors at age from the update assessment run (which can be interpreted as residuals) are shown in Figure 3.2.7 (landings), Figure 3.2.8 (discards) and Figure 3.2.9 (ScoGFS-1Q). Errors within $\pm 2$ are considered reasonable. No prediction errors against the 2010 survey data fall outside of $\pm 2$ such that no data points for this year were downweighted. Table 3.2.11 shows final parameter estimates have remained very consistent over the last four assessments.

It is important to note that the assessment is based on survey estimates of mortality with corresponding population abundance. Whilst the assumed natural mortality rate ( $\mathrm{M}=0.2$ ) is excluded from the estimates of 'fishing mortality', unallocated removals from the stock due to the fishery or other sources are not and are therefore also included in the estimates of 'fishing mortality' used in the forecast. The WG consider the mortality outputs from TSA not to represent F at age but rather estimated total mortality that can not be accounted for by the standard value used for natural mortality. These mortality estimates are here referred to as 'Z-0.2'. For management purposes, however, this combined mortality would still need to fall below the level of Flim, as higher levels of mortality over and above M are considered to have led to stock decline in the early 1980s.

Table 3.2.12 gives the TSA population numbers-at-age and Table 3.2.13 gives their associated standard errors. Estimated Z-0.2 at age is given in Table 3.2.14 and standard errors on the log of this mortality are given in Table 3.2.15. Full summary output is given in Table 3.2.16.

A summary plot for this run is shown in Figure 3.2.6. The disparity between the estimated removals compared to the supplied commercial catch data is clear. Figure 3.2.16 shows the ratio between the estimated removals and observed catch. The disparity has reduced since the largest values in 2004 and 2005 but the lower limit of the confidence intervals on the estimated removals are still above the line showing a $1: 1$ ratio.

From Figure 3.2.6 there is a noticeable long-term downward trend in recruitment although the value for the 2005 year class is the highest value since the 1996 year class and that for the 2008 year class the second highest in that same period. SSB increased in 2007 and the estimates for 2008 and 2009 are similar to that for 2007 reflecting a maturing of the 2005 year class but SSB is still well below Blim. Mean Z-0.2 is above Flim and comparable to values since 1995.

Retrospectives for the final assessment run are shown in Figure 3.2.10. This figure also shows lines at $\pm 2$ se (approximate $95 \%$ confidence limits) around the run using all years of data. Retrospective bias is small with respect to SSB. With respect to recruitment the run terminating in 2006 sits on the lower confidence limit while that terminating in 2005 falls just outside of this limit. Higher levels of Z-0.2 from the run terminating in 2005 appear untypically high but fall within the confidence limits for this metric. The confidence interval for mean Z-0.2 is wide, reflecting uncertainty in estimation of mean Z- 0.2 when that estimation is based on the age structure present in survey data.

The TSA estimated stock-recruit relationship is shown in Figure 3.2.11. It includes the data point of the 1986 year class which from inspection of Figure 3.2.11 appears an outlier.

The precautionary approach plot for this stock is given in Figure 3.2.12. It shows clearly how the stock has moved and remained in the zone indicating reduced reproductive capacity and (substituting Z-0.2 for F) unsustainable removals

## Comparison with last year's assessment

Compared to last year's assessment SSB in 2008 has been revised up from 6488 t to 6585 t while the estimate of mean Z-0.2 has risen from 0.88 to 0.91 . The estimate of recruitment in 2008 is revised up from 3.3 million to 3.9 million. The estimate of SSB in 2009 from this update assessment is 5166 t with a s.e. of 804 t . The short-term forecast from last year's assessment predicted SSB in 2009 at 5490 t which is less than one s.e. difference from the update assessment. Figure 3.2.10 shows these revisions represent comparatively small retrospective adjustments.

### 3.2.4 Short-term stock projections

A short-term projection was made using WGFRANSW following the procedure outlined in the Stock Annex.

## Estimating recruiting year-class abundance

The recruitment values (000 fish) used in the forecast are given in the following table:

| YEAR | TSA | STF |
| :--- | :--- | :--- |
| 2010 | 7062 (ScoGFS-1Q) | 7062 (ScoGFS-1Q) |
| 2011 | 5545 (Ricker) | 4697 (GM 99-08) |
| 2012 |  | 4697 (GM 99-08) |

Three-year means of the Z-0.2 estimates were taken to represent status quo mortality. At previous assessment meetings the status quo mortality was used in the intermediate and TAC years. The cod long-term management plan introduced in 2009 (Council Regulation No. 1342/2008, article 6, paragraph 4), however, directs that forecasts "assume that in the year prior to the year of application of the TAC the stock is fished with an adjustment in fishing mortality equal to the reduction in maximum allowable fishing effort that applies in that year." The TAC for 2010 remained unchanged compared to 2009 but fishing effort (kWdays) was either reduced by $25 \%$ or vessels were incorporated in schemes designed to achieve a $25 \%$ reduction in mortality. Although not considered a measure of F the status quo Z-0.2 was reduced by $25 \%$ for the intermediate year in the forecast (2010). The management options table from this first run showed SSB to be below Blim at the start of 2011. Following article 6, paragraph 2(a) of the new cod management plan status quo Z-0.2 was reduced by a further $25 \%$ for 2011 with the aim of producing more representative detailed tables.

Input data to the short-term projection are shown in Table 3.2.17. Management options from the forecast are shown in Table 3.2.18 and detailed tables of catch num-bers-at-age are shown in Table 3.2.19.

A plot of the short-term forecast is shown in Figure 3.2.13. Results from sensitivity analysis from this forecast are shown in Figure 3.2.14 and probability profiles in Figure 3.2.15. It is emphasized again that the outputs from the forecasting software include figures labeled as "H-cons" do not refer to the human consumption fishery but in the present application refer to all removals over and above the losses due to the assumed natural mortality rate of $\mathrm{M}=0.2$. These values will include estimates of unallocated fishery removals that may be due to misreporting, or additional natural mortality not encompassed by the standard value of $\mathrm{M}=0.2$. The WG recommends that these forecasts are not used to determine a future TAC using the procedure specified in Article 7 of the long-term management plan for cod, as it is not possible to determine figures for unallocated fishery removals to deduct from the forecasted total removals to calculate the TAC for 2011.

Estimates of SSB corresponding to the different levels of the Z-0.2 mortality should, however, remain appropriate. From Table 3.2.18 it can be seen that an assumption of zero removals in 2011 give an estimate of SSB in 2012 between $B_{\text {lim }}$ and $B_{\text {pa. }}$. From Figure 3.2.15 the probability of SSB in 2012 being above $\mathrm{B}_{\mathrm{pa}}$ is zero.

### 3.2.5 MSY explorations

ICES has previously defined the following PA reference points:

| Reference point | Technical basis |
| :--- | :--- |
| Bpa $=22000 \mathrm{t}$ | Previously set at 25000 t , which was considered a level at which good <br> recruitment is probable. This has since been reduced to 22000 t due to <br> an extended period of stock decline. |
| Blim $=14000 \mathrm{t}$ | Smoothed estimate of Bloss (as estimated in 1998). |
| Fpa $=0.6$ | Consistent with Bpa. |
| Flim $=0.8$ | F values above 0.8 led to stock decline in the early 1980s. |

To derive an Fmsy estimate the srmsymc package was employed. The same input data files as used for the short-term forecast were used. An alternative run using 10 year means for stock weights-at-age and mortality-at-age showed there to be little sensitivity to the averaging period used. Figure 3.2.17 shows the three stock-recruit relation-
ships fitted by the package; Ricker, Beverton-Holt and smooth hockey stick. Models were fitted using 1000 MCMC re-samples. For all three stock-recruit relationships all re-samples allowed Fmsy and Fcrash values to be determined. As such, there was no basis to reject any of the recruitment models as unsuitable for this stock. For each of the stock-recruit relationships (SRR) Figures 3.2.18 to 3.2.20 show box plots of FMSY and Fcrash together with the values of Fpa and Flim. For the Ricker and BevertonHolt SRR the estimated value of Fcrash is very close to Flim. For the smooth hockey stick SRR Fcrash is estimated between Flim and Fpa. For all three SRR the current level of Z-02 is higher than the median Fcrash value. Also the value of Fmsy is well defined and considerably lower than Fpa for all three SRR. The level of removals possible at the estimated Fmsy is poorly defined however. Circles showing the data points show values of Z-0.2 repeatedly in excess of the upper percentile for Fcrash. As expected removals and SSB have declined such that values for both are now inside confidence limits for these metrics at the estimated Z-0.2 mortality rates.

Figure 3.2.21 shows estimation of yield-per-recruit. Fmax is well defined for this stock. Comparison of Fmax to Fmsy estimated using the three SRRs (Figures 3.2..1719) shows Fmsy estimated as lower than Fmax for the Beverton-Holt model, equal for the smooth hockey stick and higher than Fmax in the Ricker model reflecting the downward slope of the stock-recruit relationship at higher SSBs.

In conclusion mortalities from removals in the range 0.17 to 0.33 are consistent with Fmsy.

### 3.2.6 Management plans

Cod in VIa is included in Council Regulation No. 1342/2008 establishing a long-term plan for cod stocks and fisheries exploiting those stocks. The plan and its evaluation by ICES were addressed by WGCSE 2009.

### 3.2.7 Uncertainties and bias in assessment and forecast

Figure 3.2.22 shows a comparison of SSB, recruitment-at-age one and mean Z-0.2 (ages2-5) estimates produced by final run assessments between this year's assessment and assessments going back to 2001.

## Landings

Since the early 1990s the most significant problem with assessment of this stock is with commercial data. Incorrect reporting of landings-species and quantity and management area-is known to have occurred and directly affects the perception of the stock. Figure 3.2.23 shows a summary plot for a run of the model where landings and discards data have been included for the years 2006-2009 (post UK 'Buyers and Sellers' and Irish 'Sales Notes' legislation, see Section 3.2.1) as well as the years prior to 1995. The commercial data has a greater precision than the survey and its inclusion in the latter years of the time-series has a strong influence on the model fit. However, the model can determine for precision but not bias. Discards at ages 3 and 4 are not being taken into account, and discard rates for theses age classes have been in excess of $80 \%$ in recent years (see 'Discards' below), nor is area misreporting. Signals of high mortality on younger ages from the survey-possibly an indicator of high predation (see 'Biological factors' below) have little influence on the model fit. Figure 3.2.23 indicates little or no misreporting in the mid to late 1990s, a period where the reliability of commercial data is already of concern. The Working Group concluded that with-
out further investigation into the discard modelling and predation, the inclusion of the 2006-2009 catch data in the assessment could not be justified.

## Effort

Commercial effort data for Division VIa from the Scottish fleets is considered very uncertain and was not used in the assessment.

## Discards

In the current set-up used for this assessment discard information is removed for the same years for which landings data is removed. The increase in discards at ages one and two since 2006 is not therefore able to influence the fit of discard parameters. Furthermore, the TSA model for VIa cod is formulated to only consider discards at ages one and two. Discards have also been recorded at older ages since 2006 and if this continues in future years, re-inclusion of discard data would probably require modification of the model to fit discard proportions across more ages.

Available discard estimates are calculated mainly from the Scottish sampling programme. The method used is to sample on a stratified basis, then raise by some auxiliary variable to, initially, total strata discards, and ultimately international discards. These estimates are prone to bias. An alternative method of raising discard data using the same raw data, and which reduces estimation bias, is being applied and tested on data from both the Northern Shelf and North Sea regions before the resulting revised data is released to assessment working groups. Data using the new method was not available for this year's assessment.

## Surveys

The survey used for this assessment changed vessel and tow duration in 1999. Although a correction has been made based on comparative tows, there will be an additional variance associated with this correction factor which will affect the survey index. The current spatial aggregation of the survey (weighted arithmetic mean) can result in hauls catching large numbers of fish having a strong influence on index values (as was the case in the ScoGFS-1Q in 2008). This in turn can cause a 'noisy' set of indices that can lead to high prediction errors from TSA (residuals form other models) and downweighting of data points. The current weighting of strata (weighting by number of valid hauls) is also not consistent between years leading to further increase in the overall estimation of survey variance. Ways of compiling the survey that can better incorporate extreme values, including new post stratification and strata weightings, are currently under investigation and are proposed for consideration at a future data compilation workshop.

## Biological factors

Assumptions on mean weight-at-length and mean maturity-at-age have remained unchanged for a long period. However, biological responses of cod in VIa as a localised species to high exploitation and low population numbers are so far unknown to the Working Group. Estimates of high predation consumption of cod relative to total stock biomass have raised concerns that natural mortality of cod at younger ages may be significantly greater than the standard value of 0.2 currently assumed and will have changed significantly over the period of the historical assessment.

## Forecasts

Short-term forecasts are sensitive to the estimation of status quo mean fishing mortality. The WG considers mortality estimates arising from an assessment heavily or wholly based on survey data are poorly estimated and therefore noisy and sensitive to survey catchability. In addition, in the case of VIa cod only one survey-series has been considered sufficiently long and self-consistent for use in assessment.

Natural mortality on cod at some or all ages is considered to have become greater than can be accommodated by the standard natural mortality figure of $M=0.2$. It is also possibly subject to a persistent upward trend. As a consequence, mortality outputs from TSA (or any model reliant on survey data) are not considered to represent a fishing mortality F at age for recent years in the time-series but rather estimates, (referred to here as 'Z-0.2'), of total mortality that cannot be accounted for by the standard value used for natural mortality. It is not possible to determine the proportion of the mortality caused by fishing and therefore not possible to partition F into landings and discard F. Until a better estimate of natural mortality can be determined short-term forecasts are only appropriate for considering the SSB corresponding to the different levels of the Z-0.2 mortality.

### 3.2.8 Recommendation for next Benchmark

The following table includes work on known problems with the current assessment package. Work prior to a benchmark assessment would also include comparison of the current assessment method with alternative methods.

| PROBLEM | SOLUTION | EXPERTISE NECESSARY | SUGGESTED TIME |
| :---: | :---: | :---: | :---: |
| Misreporting of landings. Unknown level prevents adjustment of reported catch and inclusion in assessment. | Analysis of VMS data in comparison to landings declarations to estimate the degree of area misreporting. | Requires <br> someone <br> familiar with <br> VMS analysis <br> (plus provision <br> of trip specific <br> landings <br> declarations). | Work possible in 2010. |
| Bias in discard estimates | Adoption of new discard raising methodology. | New discard raising methodology being developed as part of a PhD project. | PhD unlikely to be finished before 2012. |
| Inappropriate modelling of discards within TSA model | Revision of TSA to allow fitting of discards at higher ages. | Requires someone familiar TSA routines. | Work scheduled for 2010. |
| Variance and bias in survey index | Adoption of new aggregation methods to form final indices from haul by haul data, (combinations of new post stratification, weighting of strata and/or adoption of statistical approaches such as fitting of GAM or delta distribution models). <br> Inclusion of additional surveys (ScoQ4GFS and IGFS). ScoGFS-4Q indices to be formed in same manner as ScoGFS-1Q after conclusion of above project. <br> Addition of new survey effort and/or revision of survey design. | Work being undertaken as a Marine Scotland Science research project. | Project due for completion in 2011. <br> Comparison with existing assessment setup (single survey) possible in 2011 (after conclusion of above project). <br> Anglerfish survey records cod numbers at length, now has 5 years of data and cpue indices can be formed. Data from charter surveys in 2009 available. <br> A random stratified design for the Scottish surveys is under consideration. <br> Possible implementation in 2011. |
| Uncertainty in natural mortality (level and trend) because of unquantified predation from large and increasing seal population. | Revision of TSA to allow inclusion of different fleets, (this in turn allows estimates of age specific consumption of cod by seals to be input as if from an additional fleet). | Requires someone familiar with TSA routines. | Method for estimating age specific consumption of cod by seals presented at 2008 ICES ASC. Work to adjust TSA scheduled for 2010. |

### 3.2.9 Management considerations

The fishery is managed by a combination of TAC, area closures, technical measures and effort restrictions. These have not been effective in controlling catches which are estimated to be almost seven times greater than the TAC. Despite considerable reductions in fishing effort over the past decade, the stock structure is still truncated with few older fish present. The $25 \%$ effort reduction imposed as part of the cod long-term management plan in 2009 has not been reflected in the latest estimate of Z-0.2.

Although the UK 'Buyers and Sellers' and Irish 'Sales Notes' legislation is considered to have reduced underreporting from 2006, discard data shows increased discards at ages one and two and a change in discard practices such that fish are discarded at older ages. In 2009 discards as a percentage of catch were over $90 \%$ for ages 1 to 3 and over $80 \%$ at age 4 showing the legislation has controlled landings rather than catch. There are also reports of continued area misreporting.

Mortality estimates arising from an assessment heavily or wholly based on survey data are poorly estimated and therefore noisy and sensitive to survey catchability. In contrast, historical trends in spawning biomass and recruitment appear to be robust measures of stock dynamics.

Population estimates using the ScoGFS-1Q survey data indicated the 2005 year class to be the biggest within the last decade and the 2008 year-class to be of similar strength. Both discards at higher ages and area misreporting reduce the potential for these year classes to contribute to increases in SSB. It is important good observer coverage is conducted in Division VIa to record discard trends in future and that work is done to estimate area misreporting (comparing declared landings to VMS data).

Cod is taken in mixed demersal fisheries, and in Division VIa is now regarded as a bycatch species. To greatly reduce cod catch would likely result in having to greatly reduce harvesting of other stocks such as haddock, whiting and anglerfish. It is also important the bycatch from the Nephrops fleet is closely monitored (including discard observations). The STECF Report (STECF-SGMOS-09-05) assessing effort and catch of fishing regimes subject to fishing effort limitations shows trawl gear vessels targeting finfish (TR1 gear) to take roughly $80 \%$ of cod catch and the Nephrops fleet (TR2 gear) to take $15-20 \%$ of cod catch in ICES area VIa (Table 6.5.4.1 page 215) ICES note that the majority of TR2 vessels operating in VIa are now exempt from the effort control element of 1342/2008.

The EU cod long-term management plan, (Council Regulation No. 1342/2008) is complemented by a system of fishing effort limitation and in waters west of Scotland landings composition restrictions. For vessels of length 15 m and over operating west of a management line shown in Figure 3.2.24 effort is restricted to a lesser degree. Figure 3.2.24 also shows locations of fishing activity using TR1 gear (from VMS data) linked to cod landings in 2009, (Scottish vessels and vessels landing into Scotland). Fishing pings associated with the day entered in logbooks against cod landings have been aggregated. It can be seen a large proportion of the effort falls outside of the cod management area. If cod landings recorded in logbooks are assigned to VMS fishing pings by using the declared ICES rectangle in the logbooks the distribution of retained cod catches is as shown in Figure 3.2.25. Summing the landings gives annual totals of 55.8 tonnes inside the management line, 106.6 tonnes elsewhere in VIa in 2009. The landings composition restrictions do not restrict discards. Data provided to ICES shows discards to have increased as a proportion of catch in 2009.

Article 7 (paragraph 1) of the current management plan requires TACs to be calculated after removal of quantities of discards and fish corresponding to other sources of cod mortality caused by fishing. The current assessment of VIa cod is considered to estimate a mortality that is a combination of mortality from fishing and natural mortality not accounted for by the standard long-term input value. As such mortality from landings, discards and other causes due to fishing cannot be defined.

A report by the Sea Mammal Research unit (SMRU, 2006) gives estimates of cod consumed by grey seals to the west of Scotland and although highly uncertain, the estimates suggest predation mortality on cod is greater than can be accommodated by the standard value of natural mortality used for gadoid species in ICES Division VIa. It has not been possible using an update assessment to quantify the level of mortality caused by seal predation. This is proposed for a benchmark assessment, (see Section 3.2.9).

The values of mean Z-0.2 from the current assessment are estimates of mortality over and above M i.e. mortality from fishing plus non fishing mortality which can not be encompassed within the standard value for natural mortality. For management purposes this combined mortality would still need to fall below the level of Flim, as higher levels of mortality over and above M are considered to have led to stock decline.

Table 3.2.1. Cod in Division VIa. Official catch statistics in 1985-2009, as reported to ICES.

| Country | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 48 | 88 | 33 | 44 | 28 | - | 6 | - | 22 | 1 | 2 | + | 11 | 1 | + | + | 2 | + |
| Denmark | - | - | 4 | 1 | 3 | 2 | 2 | 3 | 2 | + | 4 | 2 | - | - | + | - | - | - |
| Faroe Islands | - | - | - | 11 | 26 | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | 7411 | 5096 | 5044 | 7669 | 3640 | 2220 | 2503 | 1957 | 3047 | 2488 | 2533 | 2253 | 956 | 714* | 842* | 236 | 391 | 208 |
| Germany | 66 | 53 | 12 | 25 | 281 | 586 | 60 | 5 | 94 | 100 | 18 | 63 | 5 | 6 | 8 | 6 | 4 | + |
| Ireland | 2564 | 1704 | 2442 | 2551 | 1642 | 1200 | 761 | 761 | 645 | 825 | 1054 | 1286 | 708 | 478 | 223 | 357 | 319 | 210 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | 2 | 1 | - | - | - | - |
| Norway | 204 | 174 | 77 | 186 | 207 | 150 | 40 | 171 | 72 | 51 | 61 | 137 | 36 | 36 | 79 | 114* | 40* | 88 |
| Spain | 28 | - | - | - | 85 | - | - | - | - | - | 16 | + | 6 | 42 | 45 | 14 | 3 | 11 |
| UK (E., W., N.I.) | 260 | 160 | 444 | 230 | 278 | 230 | 511 | 577 | 524 | 419 | 450 | 457 | 779 | 474 | 381 | 280 | 138 | 195 |
| UK (Scotland) | 8032 | 4251 | 11143 | 8465 | 9236 | 7389 | 6751 | 5543 | 6069 | 5247 | 5522 | 5382 | 4489 | 3919 | 2711 | 2057 | 1544 | 1519 |
| UK |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total landings | 18613 | 11526 | 19199 | 19182 | 15426 | 11777 | 10634 | 9017 | 10475 | 9131 | 9660 | 9580 | 6992 | 5671 | 4289 | 2767 | 2439 | 2231 |


| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |
| Denmark |  |  |  |  |  |  |  |
| Faroe Islands |  | 2 | 0 | 0.8 | 12 | 1 |  |
| France | 172 | 91 | 107 | 100.7 | 92 | 82 |  |
| Germany | + |  |  | 2 | 2 | 1 | 0 |
| Ireland | 120 | 34 | 27.9 | 18 | 70 | 58.2 | 24.4 |
| Netherlands | - |  |  |  |  |  | 0 |
| Norway | 45 | 10 | 17 | 30 | 30 | 65 | 18 |
| Spain | 3 |  |  |  |  |  |  |
| UK (E., W., N.I.) | 79 | 46 | 25 |  | 21 | 6 |  |
| UK (Scotland) | 879 | 413 | 243 |  | 260 | 232 |  |
| UK |  |  |  | 332.1 |  |  | 120 |
| Total landings | 1298 | 596 | 419.9 | 483.6 | 487 | 445.2 | 162.4 |

* Preliminary.

Table 3.2.2. Cod in Division VIa. Landings, discards and catch estimates 1978-2009, as used by the WG. Values are totals for fish over the ages 1 to $7+$.

| Year | LANDINGS | DISCARDS | Catch |
| :---: | :---: | :---: | :---: |
| 1978 | 13521 | 3678 | 17199 |
| 1979 | 16087 | 54 | 16141 |
| 1980 | 17879 | 996 | 18875 |
| $1981$ | 23866 | 520 | 24386 |
| $1982$ | 21510 | 1652 | 23162 |
| $1983$ | 21305 | 2026 | 23331 |
| $1984$ | 21271 | 635 | 21906 |
| 1985 | 18608 | 8812 | 27420 |
| 1986 | 11820 | 1201 | 13022 |
| 1987 | 18975 | 8767 | 27742 |
| $1988$ | 20413 | 1217 | 21629 |
| $1989$ | 17171 | 2833 | 20004 |
| $1990$ | 12176 | 326 | 12503 |
| $1991$ | 10926 | 917 | 11843 |
| 1992 | $9086$ | 2897 | 11983 |
| 1993 | 10315 | 192 | 10507 |
| $1994$ | 8929 | 186 | 9115 |
| 1995 | 9438 | 257 | 9696 |
| 1996 | 9425 | 87 | 9513 |
| 1997 | 7033 | 354 | 7387 |
| 1998 | 5714 | 423 | 6137 |
| $1999$ | 4201 | 98 | 4298 |
| $2000$ | 2977 | 607 | 3584 |
| $2001$ | 2347 | 224 | 2571 |
| 2002 | 2242 | 169 | 2412 |
| 2003 | 1241 | 49 | 1291 |
| 2004 | 540 | 75 | 615 |
| 2005 | 479 | 57 | 535 |
| 2006 | 463 | 478 | 940 |
| 2007 | 525 | 2104 | 2629 |
| $2008$ | 451 | 909 | 1360 |
| 2009 | 222 | 1401 | 1623 |

Table 3.2.3. Cod in Division VIa. Survey data made available to the WG. Data used in update run are highlighted in bold. For ScoGFS-1Q, numbers are standardised to catch-rate per 10 hours.

| $\begin{gathered} \text { ScoGFS- } \\ 1 Q \\ \hline \end{gathered}$ | SCOTTISH WEST COAST GROUNDFISH SURVEY |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2010 |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 0.25 |  |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |  |
| 10 | 1.5 | 23.7 | 8.6 | 13.6 | 3.9 | 2.5 | 1.2 | 1985 |
| 10 | 1.5 | 6.9 | 26.8 | 5.6 | 7.3 | 2.5 | 1.9 | 1986 |
| 10 | 57.4 | 16.2 | 15.3 | 22.8 | 3.0 | 2.8 | 0.0 | 1987 |
| 10 | 0.0 | 64.9 | 14.2 | 3.4 | 2.1 | 0.7 | 0.2 | 1988 |
| 10 | 4.5 | 7.2 | 45.1 | 8.6 | 1.9 | 0.5 | 0.8 | 1989 |
| 10 | 2.0 | 24.6 | 4.1 | 14.7 | 4.2 | 1.6 | 0.8 | 1990 |
| 10 | 4.8 | 5.4 | 17.4 | 5.2 | 13.4 | 2.8 | 0.5 | 1991 |
| 10 | 7.3 | 11.5 | 5.4 | 7.6 | 3.4 | 2.3 | 0.5 | 1992 |
| 10 | 1.7 | 38.2 | 12.7 | 1.7 | 1.4 | 1.1 | 0.0 | 1993 |
| 10 | 13.6 | 14.7 | 25.1 | 5.8 | 1.0 | 0.0 | 0.0 | 1994 |
| 10 | 6.4 | 23.8 | 14.0 | 16.5 | 1.2 | 1.9 | 0.7 | 1995 |
| 10 | 2.8 | 20.9 | 24.1 | 4.1 | 2.8 | 1.3 | 0.0 | 1996 |
| 10 | 11.1 | 7.7 | 11.6 | 7.9 | 4.2 | 4.7 | 1.0 | 1997 |
| 10 | 2.8 | 30.9 | 5.3 | 8.7 | 3.7 | 0.6 | 2.0 | 1998 |
| 10 | 1.5 | 8.2 | 8.2 | 1.4 | 3.2 | 0.5 | 0.5 | 1999 |
| 10 | 13.3 | 5.4 | 6.9 | 1.3 | 0.0 | 0.4 | 0.0 | 2000 |
| 10 | 2.7 | 18.4 | 5.7 | 13.2 | 19.5 | 1.1 | 1.6 | 2001 |
| 10 | 5.3 | 4.3 | 10.6 | 2.6 | 0.5 | 3.0 | 0.0 | 2002 |
| 10 | 2.7 | 16.7 | 2.0 | 4.7 | 1.8 | 0.7 | 0.4 | 2003 |
| 10 | 5.7 | 3.0 | 5.6 | 2.3 | 1.7 | 0.0 | 0.0 | 2004 |
| 10 | 1.3 | 1.5 | 1.2 | 0 | 0 | 0.4 | 0 | 2005 |
| 10 | 2.2 | 1.9 | 1.1 | 0.3 | 0 | 0 | 0.3 | 2006 |
| 10 | 2.1 | 18.8 | 3.4 | 1.2 | 0 | 0.6 | 0 | 2007 |
| 10 | 0.8 | 2.1 | 44.2 | 6.3 | 0.8 | 0 | 0 | 2008 |
| 10 | 1.8 | 2.6 | 2.3 | 0.4 | 0 | 0 | 0 | 2009 |
| 10 | 4.6 | 16.2 | 3.7 | 1.0 | 0.7 | 0 | 0 | 2010 |

Table 3.2.3. cont. Cod in Division VIa. Survey data made available to the WG. For IreGFS, effort is given as minutes towed, numbers are in units.

| IREGFS | IRISH GROUNDFISH SURVEY |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 1993 | 2002 |  |  |  |
| 1 | 1 | 0.75 | 0.79 |  |
| 0 | 3 |  |  | 13.0 |
| 1849 | 0.0 | 312.0 | 59.0 | 13.0 |
| 1610 | 20.0 | 999.0 | 56.0 | 69.0 |
| 1826 | 78.0 | 169.0 | 142.0 | 18.0 |
| 1765 | 0.0 | 214.0 | 89.0 | 10.0 |
| 1581 | 6.0 | 565.0 | 31.0 | 6.0 |
| 1639 | 0.0 | 83.0 | 14.0 | 3.0 |
| 1564 | 0.0 | 24.0 | 4.0 | 1.0 |
| 1556 | 0.0 | 124.0 | 28.0 | 2.0 |
| 755 | 3.0 | 82.0 | 2.2 | 1.2 |
| 798 | 0.0 | 50.6 |  |  |

Table 3.2.3. cont. Cod in Division VIa. Survey data made available to the WG. For ScoGFS-4Q, numbers are standardised to catch-rate per 10 hours. " + " indicates value less than 0.5 after standardising.

| ScoGFS-4Q |  | QUARTER 4 SCOTTISH GROUND FISH SURVEY |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 2009 |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 1.00 |  |  |  |  |  |  |  |
| 0 | 8 |  |  |  |  |  |  |  |  |  |
| 10 | 0 | 1 | 14 | 5 | 3 | 1 | 0 | 0 | 0 | 1996 |
| 10 | 1 | 11 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 1997 |
| 10 | + | 15 | 9 | 1 | 0 | 0 | 0 | 0 | 0 | 1998 |
| 10 | 2 | 4 | 6 | 9 | 1 | 0 | 0 | 0 | 0 | 1999 |
| 10 | 0 | 16 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 2000 |
| 10 | 1 | 2 | 9 | 1 | 1 | 0 | 0 | 0 | 0 | 2001 |
| 10 | 1 | 10 | 3 | 7 | 1 | 0 | 0 | 0 | 0 | 2002 |
| 10 | 1 | 2 | 11 | 3 | 1 | 0 | 0 | 0 | 0 | 2003 |
| 10 | 0 | 5 | 4 | 0 | + | 0 | 0 | 0 | 0 | 2004 |
| 10 | + | 2 | 3 | 0 | 1 | + | 0 | 0 | 0 | 2005 |
| 10 | 0 | 17 | 6 | 1 | 1 | 0 | 0 | 0 | 0 | 2006 |
| 10 | 0 | 12.0 | 20.0 | 1.3 | 0.6 | 0 | 0.3 | 0 | 0 | 2007 |
| 10 | 2 | 8 | 5 | 7 | 1 | 0 | 0 | 0 | 0 | 2008 |
| 10 | 2 | 14 | 4 | 1 | 1 | + | 0 | 0 | 0 | 2009 |

Table 3.2.3. cont. Cod in Division VIa. Survey data made available to the WG. For IGFS, effort is given as minutes towed, numbers are in units. Values for 2007 are revised compared to last year's assessment.

| IGFS | IRISH WEST CoASt GROUNDFISH |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 2009 |  |  |  |  |  |
| 1 | 1 |  | 0.92 |  |  |  |
| 0 | 4 | 10 | 11 | 0 | 0 | 2003 |
| 1127 | 0 | 24 | 10 | 1 | 0 | 2004 |
| 1200 | 0 | 13 | 7 | 0 | 2 | 2005 |
| 960 | 63 | 95 | 12 | 0 | 0 | 2006 |
| 1510 | 0 | 161 | 12 | 0 | 1 | 2007 |
| 1173 | 0 | 23 | 24 | 4 | 0 | 2008 |
| 1135 | 0 | 75 | 4 | 5 | 0 | 2009 |
| 1378 | 1 |  |  |  |  |  |

Table 3.2.4. Cod in Division VIa. Landings-at-age (thousands).

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1966 | 384 | 2883 | 629 | 999 | 825 | 78 | 52 |
| 1967 | 261 | 2571 | 3705 | 670 | 442 | 264 | 67 |
| 1968 | 333 | 1364 | 3289 | 1838 | 215 | 171 | 151 |
| 1969 | 64 | 1974 | 1332 | 1943 | 759 | 149 | 170 |
| 1970 | 256 | 1176 | 1638 | 571 | 476 | 153 | 74 |
| 1971 | 254 | 1903 | 550 | 841 | 240 | 201 | 95 |
| 1972 | 735 | 2891 | 1591 | 409 | 501 | 108 | 110 |
| 1973 | 1015 | 1524 | 1442 | 583 | 161 | 193 | 104 |
| 1974 | 843 | 2318 | 778 | 1068 | 288 | 72 | 102 |
| 1975 | 1207 | 1898 | 1187 | 533 | 325 | 90 | 35 |
| 1976 | 970 | 3682 | 1467 | 638 | 256 | 215 | 56 |
| 1977 | 1265 | 1314 | 1639 | 624 | 269 | 87 | 79 |
| 1978 | 723 | 1761 | 999 | 695 | 286 | 97 | 75 |
| 1979 | 929 | 1612 | 2125 | 682 | 342 | 134 | 69 |
| 1980 | 1195 | 3294 | 2001 | 796 | 191 | 77 | 37 |
| 1981 | 461 | 7016 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 1827 | 1673 | 3206 | 1189 | 367 | 111 | 33 |
| 1983 | 2335 | 4515 | 1118 | 1400 | 468 | 148 | 60 |
| 1984 | 2143 | 2360 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 1355 | 5069 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 792 | 1486 | 2055 | 411 | 191 | 40 | 30 |
| 1987 | 7873 | 4837 | 988 | 905 | 137 | 56 | 26 |
| 1988 | 1008 | 8336 | 2193 | 278 | 210 | 39 | 20 |
| 1989 | 2017 | 1082 | 3858 | 709 | 113 | 69 | 33 |
| 1990 | 513 | 4024 | 432 | 924 | 170 | 23 | 11 |
| 1991 | 1518 | 1728 | 1805 | 188 | 266 | 70 | 23 |
| 1992 | 1407 | 1868 | 575 | 720 | 69 | 58 | 24 |
| 1993 | 328 | 3596 | 1050 | 131 | 183 | 24 | 36 |
| 1994 | 942 | 1207 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 753 | 2750 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 341 | 2331 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | 1414 | 1067 | 989 | 281 | 66 | 62 | 7 |
| 1998 | 310 | 3318 | 293 | 174 | 57 | 16 | 9 |
| 1999 | 132 | 884 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 765 | 532 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 96 | 1241 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 337 | 340 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 62 | 516 | 85 | 107 | 6 | 2 | 1 |
| 2004 | 44 | 92 | 85 | 11 | 26 | 2 | 1 |
| 2005 | 31 | 121 | 43 | 37 | 7 | 6 | 0.5 |
| 2006 | 17 | 91 | 72 | 21 | 13 | 2 | 1 |
| 2007 | 5 | 165 | 62 | 33 | 3 | 3 | 2 |
| 2008 | 0.07 | 27 | 88 | 16 | 10 | 1 | 2 |
| 2009 | 2 | 10 | 9 | 30 | 4 | 1 | 0.1 |

Table 3.2.5. Cod in Division VIa. Mean weight-at-age in landings (kg).

| Year | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1966 | 0.730 | 1.466 | 3.474 | 5.240 | 4.868 | 8.711 | 9.250 |
| 1967 | 0.681 | 1.470 | 2.906 | 4.560 | 6.116 | 7.394 | 8.058 |
| 1968 | 0.745 | 1.776 | 2.766 | 4.721 | 6.304 | 7.510 | 8.278 |
| 1969 | 0.860 | 1.284 | 2.821 | 4.259 | 6.169 | 6.374 | 7.928 |
| 1970 | 0.595 | 0.955 | 2.533 | 4.678 | 6.016 | 7.120 | 8.190 |
| 1971 | 0.674 | 1.046 | 2.536 | 4.167 | 6.023 | 6.835 | 8.100 |
| 1972 | 0.609 | 1.192 | 2.586 | 4.417 | 6.226 | 7.585 | 8.538 |
| 1973 | 0.597 | 1.181 | 2.784 | 4.601 | 5.625 | 7.049 | 8.611 |
| 1974 | 0.611 | 1.103 | 2.834 | 4.750 | 6.144 | 7.729 | 9.339 |
| 1975 | 0.603 | 1.369 | 3.078 | 5.302 | 6.846 | 8.572 | 10.328 |
| 1976 | 0.616 | 1.397 | 3.161 | 5.005 | 6.290 | 8.017 | 9.001 |
| 1977 | 0.629 | 1.160 | 2.605 | 4.715 | 6.269 | 7.525 | 9.511 |
| 1978 | 0.630 | 1.373 | 3.389 | 5.262 | 7.096 | 8.686 | 9.857 |
| 1979 | 0.693 | 1.373 | 2.828 | 4.853 | 6.433 | 7.784 | 9.636 |
| 1980 | 0.624 | 1.375 | 3.002 | 5.277 | 7.422 | 8.251 | 9.331 |
| 1981 | 0.550 | 1.166 | 2.839 | 4.923 | 7.518 | 9.314 | 10.328 |
| 1982 | 0.692 | 1.468 | 2.737 | 4.749 | 6.113 | 7.227 | 9.856 |
| 1983 | 0.583 | 1.265 | 2.995 | 4.398 | 6.305 | 8.084 | 9.744 |
| 1984 | 0.735 | 1.402 | 3.168 | 5.375 | 6.601 | 8.606 | 10.350 |
| 1985 | 0.628 | 1.183 | 2.597 | 4.892 | 6.872 | 8.344 | 9.766 |
| 1986 | 0.710 | 1.211 | 2.785 | 4.655 | 6.336 | 8.283 | 9.441 |
| 1987 | 0.531 | 1.312 | 2.783 | 4.574 | 6.161 | 7.989 | 10.062 |
| 1988 | 0.806 | 1.182 | 2.886 | 5.145 | 6.993 | 8.204 | 9.803 |
| 1989 | 0.704 | 1.298 | 2.425 | 4.737 | 7.027 | 7.520 | 9.594 |
| 1990 | 0.613 | 1.275 | 2.815 | 4.314 | 7.021 | 9.027 | 11.671 |
| 1991 | 0.640 | 1.095 | 2.618 | 4.346 | 6.475 | 8.134 | 10.076 |
| 1992 | 0.686 | 1.293 | 2.607 | 4.268 | 6.190 | 7.844 | 10.598 |
| 1993 | 0.775 | 1.316 | 2.940 | 4.646 | 6.244 | 7.802 | 8.409 |
| 1994 | 0.644 | 1.292 | 2.899 | 4.710 | 6.389 | 8.423 | 8.409 |
| 1995 | 0.606 | 1.148 | 2.857 | 4.956 | 6.771 | 8.539 | 9.505 |
| 1996 | 0.667 | 1.221 | 2.738 | 5.056 | 6.892 | 8.088 | 10.759 |
| 1997 | 0.595 | 1.210 | 2.571 | 4.805 | 6.952 | 7.821 | 9.630 |
| 1998 | 0.605 | 1.061 | 2.264 | 4.506 | 6.104 | 8.017 | 9.612 |
| 1999 | 0.691 | 1.039 | 2.194 | 4.688 | 6.486 | 8.252 | 9.439 |
| 2000 | 0.689 | 1.261 | 2.457 | 4.126 | 6.666 | 7.917 | 8.392 |
| 2001 | 0.654 | 0.988 | 2.679 | 4.568 | 5.860 | 7.741 | 9.386 |
| 2002 | 0.668 | 1.140 | 2.330 | 4.841 | 6.175 | 7.192 | 9.548 |
| 2003 | 0.671 | 1.016 | 2.312 | 3.854 | 6.220 | 8.075 | 8.839 |
| 2004 | 0.609 | 1.027 | 2.194 | 4.396 | 6.003 | 8.258 | 9.678 |
| 2005 | 0.776 | 1.172 | 2.624 | 4.118 | 4.908 | 6.753 | 10.240 |
| 2006 | 0.656 | 1.169 | 2.236 | 3.822 | 6.172 | 7.796 | 11.1 |
| 2007 | 0.476 | 0.976 | 2.512 | 4.285 | 6.491 | 7.733 | 8.810 |
| 2008 | 0.557 | 1.195 | 2.943 | 4.775 | 6.329 | 7.957 | 8.471 |
| 2009 | 1.048 | 1.960 | 2.916 | 4.743 | 5.853 | 8.171 | 8.646 |

Table 3.2.6. Cod in Division VIa. Discard dataset from Scottish \& Irish sampling programmes, ages 1-7, years 1978-2008. Data from 1978-2001 raised from Scottish sampling only; later data raised from Scottish sampling and Irish sampling when available ( $2004 \& 2005$ to date).

| DIsCards at age (thousands) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Age |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1978 | 8904 | 1203 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 11 | 119 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 2758 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 289 | 1475 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 5264 | 2 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 7371 | 1005 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 2117 | 10 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 43508 | 3122 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 4483 | 10 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 52582 | 159 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 714 | 3256 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 8443 | 25 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1835 | 158 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 3255 | 319 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 12498 | 143 | 2 | 0 | 0 | 0 | 0 |
| 1993 | 595 | 51 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 773 | 2 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 1111 | 126 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 233 | 86 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 1074 | 27 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 472 | 837 | 3 | 0 | 0 | 0 | 0 |
| 1999 | 283 | 16 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 2081 | 53 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 216 | 373 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 508 | 32 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 77 | 38 | 8 | 0 | 0 | 0 | 0 |
| 2004 | 232 | 21 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 108 | 20 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 1242 | 48 | 25 | 2 | 3 | 1 | 0.1 |
| 2007 | 627 | 1651 | 56 | 42 | 3 | 3 | 0 |
| 2008 | 89 | 133 | 368 | 1 | 0 | 0 | 0 |
| 2009 | 883 | 219 | 160 | 138 | 0 | 7 | 0 |

Table 3.2.7. Cod in Division VIa. Discard dataset from Scottish \& Irish sampling programmes, ages 1-7, years 1978-2006. Data from 1978-2001 raised from Scottish sampling only; later data raised from Scottish sampling and Irish sampling when available (2004 \& 2005 to date).

| Mean weight-AT-AGE in discards (kG) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Age |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1978 | 0.37 | 0.321 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0.276 | 0.43 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0.361 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0.135 | 0.326 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0.314 | 0.392 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0.223 | 0.374 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 0.298 | 0.435 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 0.178 | 0.346 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 0.267 | 0.305 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0.166 | 0.37 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0.296 | 0.283 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0.332 | 0.59 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0.132 | 0.454 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0.245 | 0.351 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0.22 | 1.03 | 2.382 | 0 | 0 | 0 | 0 |
| 1993 | 0.239 | 0.812 | 3.723 | 0 | 0 | 0 | 0 |
| 1994 | 0.24 | 0.365 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0.203 | 0.256 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0.226 | 0.389 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0.321 | 0.328 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0.23 | 0.367 | 0.59 | 0 | 0 | 0 | 0 |
| 1999 | 0.294 | 0.299 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0.28 | 0.421 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0.248 | 0.417 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0.263 | 1.021 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0.272 | 0.57 | 0.39 | 0 | 0 | 0 | 0 |
| 2004 | 0.258 | 0.581 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0.285 | 0.501 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0.259 | 1.291 | 2.649 | 3.499 | 6.24 | 5.581 | 11.122 |
| 2007 | 0.198 | 0.940 | 3.016 | 4.453 | 5.018 | 10.627 | 0 |
| 2008 | 0.220 | 0.976 | 2.046 | 4.047 | 7.937 | 0 | 0 |
| 2009 | 0.261 | 1.312 | 2.248 | 3.324 | 0 | 6.448 | 0 |

Table 3.2.8. Cod in Division VIa. Total catch-at-age (thousands).

| Year | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 9627 | 2965 | 999 | 695 | 286 | 97 | 75 |
| 1979 | 940 | 1731 | 2125 | 682 | 342 | 134 | 69 |
| 1980 | 3953 | 3294 | 2001 | 796 | 191 | 77 | 37 |
| 1981 | 749 | 8491 | 3220 | 904 | 182 | 29 | 20 |
| 1982 | 7091 | 1676 | 3206 | 1189 | 367 | 111 | 33 |
| $1983$ | 9706 | 5520 | 1118 | 1400 | 468 | 148 | 60 |
| $1984$ | 4260 | 2371 | 2564 | 448 | 555 | 185 | 59 |
| 1985 | 44863 | 8191 | 1269 | 1091 | 140 | 167 | 79 |
| 1986 | 5275 | 1495 | 2055 | 411 | 191 | 40 | 30 |
| $1987$ | 60456 | 4996 | 988 | 905 | 137 | 56 | 26 |
| $1988$ | 1722 | 11592 | 2193 | 278 | 210 | 39 | 20 |
| $1989$ | 10459 | 1107 | 3858 | 709 | 113 | 69 | 33 |
| $1990$ | $2348$ | 4182 | 432 | 924 | 170 | 23 | 11 |
| $1991$ | 4773 | 2047 | 1805 | 188 | 266 | 70 | 23 |
| 1992 | 13905 | 2011 | 577 | 720 | 69 | 58 | 24 |
| 1993 | 923 | 3647 | 1050 | 131 | 183 | 24 | 36 |
| 1994 | 1715 | 1209 | 1545 | 280 | 56 | 51 | 20 |
| 1995 | 1864 | 2877 | 700 | 630 | 70 | 15 | 11 |
| 1996 | 574 | 2417 | 1210 | 247 | 204 | 31 | 13 |
| 1997 | $2488$ | 1094 | 989 | 281 | 66 | 62 | 7 |
| $1998$ | 783 | 4155 | 296 | 174 | 57 | 16 | 9 |
| $1999$ | 415 | 900 | 1047 | 64 | 48 | 24 | 9 |
| 2000 | 2846 | 585 | 211 | 231 | 15 | 12 | 13 |
| 2001 | 312 | 1614 | 155 | 63 | 52 | 3 | 4 |
| 2002 | 845 | 372 | 522 | 41 | 13 | 14 | 4 |
| 2003 | 139 | 554 | 93 | 107 | 6 | 2 | 1 |
| 2004 | 267 | 113 | 85 | 11 | 26 | 2 | 1 |
| 2005 | 139 | 141 | 43 | 37 | 7 | 6 | 0.5 |
| 2006 | 1259 | 139 | 97 | 23 | 15 | 2 | 1 |
| 2007 | 632 | 1816 | 118 | 75 | 5 | 7 | 2 |
| 2008 | 89 | 160 | 456 | 18 | 10 | 1 | 2 |
| 2009 | 885 | 229 | 168 | 168 | 4 | 8 | 0.1 |

Table 3.2.9. Cod in Division VIa. Mean weight-at-age (kg) in total catch.

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 0.389 | 0.946 | 3.389 | 5.262 | 7.096 | 8.686 | 9.857 |
| 1979 | 0.688 | 1.308 | 2.828 | 4.853 | 6.433 | 7.784 | 9.636 |
| 1980 | 0.440 | 1.375 | 3.002 | 5.277 | 7.422 | 8.251 | 9.331 |
| 1981 | 0.390 | 1.020 | 2.839 | 4.923 | 7.518 | 9.314 | 10.328 |
| 1982 | 0.411 | 1.467 | 2.737 | 4.749 | 6.113 | 7.227 | 9.856 |
| $1983$ | $0.310$ | 1.103 | 2.995 | 4.398 | 6.305 | 8.084 | 9.744 |
| $1984$ | 0.518 | 1.398 | 3.168 | 5.375 | 6.601 | 8.606 | 10.350 |
| 1985 | 0.191 | 0.864 | 2.597 | 4.892 | 6.872 | 8.344 | 9.766 |
| 1986 | 0.334 | 1.205 | 2.785 | 4.655 | 6.336 | 8.283 | 9.441 |
| $1987$ | 0.213 | 1.282 | 2.783 | 4.574 | 6.161 | 7.989 | 10.062 |
| $1988$ | $0.595$ | 0.929 | 2.886 | 5.145 | 6.993 | 8.204 | 9.803 |
| $1989$ | $0.404$ | 1.282 | 2.425 | 4.737 | 7.027 | 7.520 | 9.594 |
| $1990$ | $0.237$ | 1.244 | 2.815 | 4.314 | 7.021 | 9.027 | 11.671 |
| $1991$ | 0.371 | 0.979 | 2.618 | 4.346 | 6.475 | 8.134 | 10.076 |
| 1992 | 0.267 | 1.274 | 2.606 | 4.268 | 6.190 | 7.844 | 10.598 |
| 1993 | 0.430 | 1.309 | 2.940 | 4.646 | 6.244 | 7.802 | 8.409 |
| 1994 | 0.462 | 1.291 | 2.899 | 4.710 | 6.389 | 8.423 | 8.409 |
| 1995 | $0.365$ | 1.109 | 2.857 | 4.956 | 6.771 | 8.539 | 9.505 |
| 1996 | $0.487$ | 1.191 | 2.738 | 5.056 | 6.892 | 8.088 | 10.759 |
| $1997$ | $0.477$ | 1.188 | 2.571 | 4.805 | 6.952 | 7.821 | 9.630 |
| $1998$ | $0.379$ | 0.921 | 2.248 | 4.506 | 6.104 | 8.017 | 9.612 |
| $1999$ | 0.420 | 1.025 | 2.194 | 4.688 | 6.486 | 8.252 | 9.439 |
| 2000 | 0.390 | 1.186 | 2.457 | 4.126 | 6.666 | 7.917 | 8.392 |
| 2001 | 0.372 | 0.856 | 2.679 | 4.568 | 5.860 | 7.741 | 9.386 |
| 2002 | 0.424 | 1.130 | 2.330 | 4.841 | 6.175 | 7.192 | 9.548 |
| 2003 | 0.450 | 0.986 | 2.15 | 3.854 | 6.220 | 8.075 | 8.839 |
| 2004 | 0.314 | 0.945 | 2.194 | 4.396 | 6.003 | 8.258 | 9.678 |
| 2005 | 0.395 | 1.078 | 2.624 | 4.118 | 4.908 | 6.753 | 10.240 |
| 2006 | 0.264 | 1.211 | 2.341 | 3.797 | 6.184 | 7.031 | 11.103 |
| 2007 | 0.200 | 0.943 | 2.752 | 4.380 | 5.729 | 9.166 | 8.810 |
| 2008 | 0.220 | 1.013 | 2.219 | 4.731 | 6.371 | 7.957 | 8.471 |
| 2009 | 0.262 | 1.340 | 2.283 | 3.577 | 5.853 | 6.654 | 8.646 |

Table 3.2.10. Cod in Division VIa. TSA parameter settings for the assessment run.

| Parameter | Setting | Justification |
| :---: | :---: | :---: |
| Age of full selection. | $\mathrm{am}=4$ | Based on inspection of previous XSA runs. |
| Multipliers on variance matrices of measurements. | Blandings $(a)=2$ for ages 6, 7+ <br> Bsurvey $(a)=2$ for age 1, 5 , 6 | Allows extra measurement variability for poorly-sampled ages. |
| Multipliers on variances for fishing mortality estimates. | $\mathrm{H}(1)=4$ | Allows for more variable fishing mortalities for age 1 fish. |
| Downweighting of particular data points (implemented by multiplying the relevant $q$ by 9) | Landings: age 2 in 1981 and 1987, age 7 in 1989. <br> Discards: age 1 in 1985 and 1992, age 2 in 1998. <br> Survey: age 1 in 2000, age 2 in 1993, age 6 in 1995. Ages 4, 5, 6 in 2001 (the latter are from a single large haul, 24 fish $>75 \mathrm{~cm}$ in 30 mins.). Age 3 in 2008 (large haul near 4W line) | Large values indicated by exploratory prediction error plots. |
| Discards | Discards are allowed to ev Ages 1 and 2 are modelled | e over time constrained by a trend. dependently. |
| Recruitment. | Modelled by a Ricker model independent and normally where $S$ is the spawning stock year. To allow recruitment recruitment, a constant coeff | with numbers-at-age 1 assumed to be stributed with mean $\eta 1 S \exp (-\eta 2 S)$, $k$ biomass at the start of the previous variability to increase with mean cient of variation is assumed. |
| Large year classes. | The 1986 year class was large not well modelled by the R $N(1,1980)$ is taken to be no $5 \eta 1 S \exp (-\eta 2 S)$. The facto maximum recruitment to $m$ VIa cod, haddock, and whi The coefficient of variation | , and recruitment at age 1 in 1987 is ker recruitment model. Instead, ally distributed with mean of 5 was chosen by comparing dian recruitment from 1966-1996 for g in turn using previous XSA runs. again assumed to be constant. |

Table 3.2.11. Cod in Division VIa. TSA parameter estimates for 2002-2004, 2006-2009 assessments and final assessment presented this year. No final assessment using TSA was conducted in 2005. Run 3 from 2004 used a similar approach to this year's assessment.

| Parameter | Notation | Description | 2002 WG | 2003 WG | $\begin{gathered} 2004 \text { WG } \\ \text { RUN } 3 \end{gathered}$ | 2006 WG | 2007 WG | 2008 WG | 2009WG | 2010WG |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Initial fishing mortality | $F(1,1978)$ | Fishing mortality-at-age $a$ in year $y$ | 0.03 | 0.64 | 0.64 | 0.6378 | 0.6337 | 0.6366 | 0.6373 | 0.6334 |
|  | $F(2,1978)$ |  | 0.25 | 0.62 | 0.57 | 0.5333 | 0.5889 | 0.5803 | 0.5797 | 0.5853 |
|  | $F(4,1978)$ |  | 0.67 | 0.82 | 0.66 | 0.5743 | 0.6879 | 0.5888 | 0.5886 | 0.5955 |
| Survey selectivities | $\Phi(1)$ | Survey selectivity-at-age $a$ | 0.83 | 0.33 | 0.47 | 0.6275 | 0.5425 | 0.4746 | 0.4809 | 0.4791 |
|  | $\Phi(2)$ |  | 4.41 | 1.98 | 3.19 | 3.5857 | 3.7292 | 3.2855 | 3.3317 | 3.3463 |
|  | $\Phi(4)$ |  | 18.28 | 10.65 | 14.92 | 15.9096 | 14.1997 | 14.0472 | 13.7891 | 13.6507 |
| Fishing mortality standard deviations | $\sigma_{F}$ | Transitory changes in overall fishing mortality | 0.10 | 0.04 | 0.07 | 0.0947 | 0.0741 | 0.0846 | 0.0850 | 0.0834 |
|  | $\sigma u$ | Persistent changes in selection (age effect in F) | 0.10 | 0.06 | 0.03 | 0.0242 | 0.0507 | 0.00 | 0.00 | 0.0057 |
|  | $\sigma V$ | Transitory changes in the year effect in fishing mortality | 0.00 | 0.07 | 0.10 | 0.0844 | 0.0984 | 0.1120 | 0.1117 | 0.1144 |
|  | $\sigma Y$ | Persistent changes in the year effect in fishing mortality | 0.16 | 0.07 | 0.00 | 0.0425 | 0.00 | 0.00 | 0.00 | 0.00 |
| Survey catchability standard deviations | $\sigma_{\Omega}$ | Transitory changes in survey catchability | 0.24 | 0.00 | 0.00 | 0.1224 | 0.2374 | 0.2276 | 0.2498 | 0.2275 |
|  | $\sigma_{\beta}$ | Persistent changes in survey catchability | 0.00 | 0.45 | 0.00 (f) | 0.00 (f) | 0.00 (f) | 0.00 (f) | 0.00(f) | 0.00(f) |


| Parameter | Notation | Description | 2002 WG | 2003 WG | $\begin{aligned} & 2004 \text { WG } \\ & \text { RuN } 3 \end{aligned}$ | 2006 WG | 2007 WG | 2008 WG | 2009WG | 2010WG |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Measurement standard deviations | Olandings | Standard error of landings-atage data | 0.12 | 0.13 | 0.10 | 0.0935 | 0.0891 | 0.0892 | 0.0889 | 0.0897 |
|  | $\sigma$ discards | Standard error of discards-at-age data | $\mathrm{n} / \mathrm{a}$ | 0.94 | 1.42 | 1.2669 | 1.367 | 1.3756 | 1.3681 | 1.3819 |
|  | $\sigma_{\text {survey }}$ | Standard error of survey data | 0.36 | 0.56 | 0.35 | 0.3887 | 0.364 | 0.3875 | 0.3930 | 0.3926 |
| Discards | $\sigma$ Oogit p | Transitory trends in discarding | n/a | 0.30 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  | $\sigma_{\text {persistent }}$ | Persistent trends in discarding | n/a | 0.16 | 0.68 | 0.5735 | 0.6742 | 0.7032 | 0.6959 | 0.7112 |
| Recruitment | $\eta_{1}$ | Ricker parameter (slope at the origin) | 0.82 | 0.62 | 0.80 | 0.6584 | 0.7882 | 0.9634 | 0.8913 | 1.0233 |
|  | $\eta_{2}$ | Ricker parameter (curve dome occurs at $1 / \eta^{2}$ ) | 0.03 | 0.003 | 0.01 | 0.0049 | 0.0124 | 0.0203 | 0.0177 | 0.0223 |
|  | ${ }^{\text {c }}$ rec | Coefficient of variation of recruitment data | 0.36 | 0.56 | 0.49 | 0.4184 | 0.5116 | 0.5627 | 0.5530 | 0.5671 |

Tabele 3.2.12. Cod in Division VIa. TSA population numbers-at-age (millions).

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 20.5672 | 9.4801 | 2.5819 | 1.4145 | 0.5274 | 0.1602 | 0.1300 |
| 1979 | 28.3269 | 10.1966 | 4.2357 | 1.1249 | 0.5277 | 0.1860 | 0.1000 |
| 1980 | 31.0962 | 13.7576 | 4.3554 | 1.3395 | 0.2795 | 0.1205 | 0.0600 |
| 1981 | 10.5034 | 16.3109 | 6.1920 | 1.8093 | 0.4951 | 0.1000 | 0.0655 |
| 1982 | 25.7117 | 5.1062 | 6.8386 | 2.3878 | 0.6802 | 0.1912 | 0.0598 |
| $1983$ | 15.4779 | 12.0399 | 2.1641 | 2.5890 | 0.8554 | 0.2395 | 0.0888 |
| 1984 | 24.0547 | 6.0868 | 4.5504 | 0.7565 | 0.8460 | 0.2782 | 0.1023 |
| 1985 | 12.3570 | 12.0726 | 2.2218 | 1.4568 | 0.2280 | 0.2285 | 0.1106 |
| $1986$ | 19.1694 | 4.2428 | 3.8769 | 0.6899 | 0.3258 | 0.0632 | 0.0784 |
| $1987$ | 59.8868 | 9.8640 | 1.7519 | 1.3639 | 0.2271 | 0.1029 | 0.0464 |
| $1988$ | 6.0555 | 16.8212 | 3.6523 | 0.5509 | 0.3553 | 0.0664 | 0.0422 |
| $1989$ | 19.7253 | 2.4807 | 5.4805 | 1.1630 | 0.1862 | 0.1076 | 0.0346 |
| $1990$ | 6.3566 | 8.7653 | 0.9425 | 1.4928 | 0.3405 | 0.0555 | 0.0404 |
| 1991 | 11.1191 | 2.9277 | 3.4216 | 0.3637 | 0.4826 | 0.1211 | 0.0348 |
| 1992 | 17.3426 | 4.5392 | 0.9635 | 1.1371 | 0.1240 | 0.1482 | 0.0478 |
| $1993$ | 7.1597 | 8.1772 | 1.8265 | 0.3055 | 0.3407 | 0.0417 | 0.0664 |
| $1994$ | 15.0998 | 3.3037 | 3.2895 | 0.6068 | 0.1089 | 0.1130 | 0.0367 |
| 1995 | 12.6504 | 7.5079 | 1.4766 | 1.3010 | 0.2290 | 0.0403 | 0.0559 |
| $1996$ | 4.9338 | 5.9165 | 3.0450 | 0.5375 | 0.4540 | 0.0792 | 0.0332 |
| $1997$ | 17.6711 | 2.0668 | 2.2954 | 1.0416 | 0.1786 | 0.1524 | 0.0372 |
| $1998$ | 8.7168 | 7.9998 | 0.7721 | 0.7837 | 0.3351 | 0.0581 | 0.0617 |
| 1999 | 4.9317 | 3.8497 | 3.0104 | 0.2524 | 0.2497 | 0.1062 | 0.0382 |
| 2000 | 10.2059 | 2.1109 | 1.4767 | 1.0174 | 0.0805 | 0.0799 | 0.0463 |
| $2001$ | 3.2490 | 4.6891 | 0.8363 | 0.5251 | 0.3455 | 0.0271 | 0.0425 |
| 2002 | 8.7484 | 1.3206 | 1.7704 | 0.2796 | 0.1677 | 0.1137 | 0.0224 |
| $2003$ | 1.6749 | 3.8883 | 0.4857 | 0.5895 | 0.0892 | 0.0538 | 0.0437 |
| 2004 | 3.8522 | 0.5549 | 1.4034 | 0.1579 | 0.1818 | 0.0271 | 0.0302 |
| 2005 | 5.3315 | 1.3827 | 0.1543 | 0.4346 | 0.0457 | 0.0546 | 0.0172 |
| 2006 | 11.0528 | 2.3017 | 0.4484 | 0.0359 | 0.1298 | 0.0134 | 0.0215 |
| 2007 | 2.4881 | 5.0325 | 0.9110 | 0.1568 | 0.0118 | 0.0434 | 0.0116 |
| 2008 | 3.8602 | 1.1427 | 2.0370 | 0.3307 | 0.0534 | 0.0040 | 0.0188 |
| 2009 | 10.3900 | 1.6939 | 0.4316 | 0.6614 | 0.1028 | 0.0164 | 0.0071 |
| 2010* | 7.0617 | 4.6992 | 0.6510 | 0.1447 | 0.2094 | 0.0325 | 0.0075 |
| 2011* | 5.5452 | 3.2290 | 1.8545 | 0.2279 | 0.0477 | 0.0692 | 0.0132 |
| GM(78-09) | 10.2398 | 4.5723 | 1.8231 | 0.6630 | 0.2184 | 0.0734 | 0.0427 |

[^0]Table 3.2.13. Cod in Division VIa. Standard errors on TSA population numbers-at-age (millions).

| Year | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 2.9004 | 0.542 | 0.1236 | 0.0845 | 0.0511 | 0.0299 | 0.022 |
| 1979 | 2.1581 | 0.5551 | 0.1765 | 0.0605 | 0.0448 | 0.0306 | 0.0192 |
| 1980 | 2.5842 | 0.787 | 0.2327 | 0.0973 | 0.0313 | 0.027 | 0.0197 |
| 1981 | 1.1827 | 1.2187 | 0.3394 | 0.1007 | 0.0375 | 0.0138 | 0.0126 |
| 1982 | 2.1817 | 0.3662 | 0.3827 | 0.1356 | 0.038 | 0.0147 | 0.0046 |
| 1983 | 1.5714 | 0.9024 | 0.1161 | 0.1641 | 0.0652 | 0.0245 | 0.0089 |
| 1984 | 1.7581 | 0.5346 | 0.2854 | 0.051 | 0.0713 | 0.0351 | 0.0135 |
| 1985 | 1.4787 | 0.8001 | 0.1501 | 0.114 | 0.0232 | 0.0374 | 0.0187 |
| 1986 | 1.4828 | 0.3271 | 0.237 | 0.0522 | 0.041 | 0.0114 | 0.0174 |
| 1987 | 10.0221 | 0.6542 | 0.1017 | 0.0959 | 0.0219 | 0.019 | 0.0091 |
| 1988 | 1.1223 | 1.5986 | 0.1928 | 0.0376 | 0.0356 | 0.0109 | 0.0088 |
| 1989 | 2.0397 | 0.1831 | 0.4723 | 0.076 | 0.0142 | 0.0158 | 0.0065 |
| 1990 | 1.1508 | 0.4923 | 0.0528 | 0.1315 | 0.0286 | 0.0071 | 0.0068 |
| 1991 | 1.5709 | 0.2207 | 0.1963 | 0.0199 | 0.0426 | 0.0131 | 0.004 |
| 1992 | 1.645 | 0.3177 | 0.0701 | 0.0769 | 0.0088 | 0.0191 | 0.0063 |
| 1993 | 1.0106 | 0.5082 | 0.1268 | 0.0254 | 0.0326 | 0.0047 | 0.0083 |
| 1994 | 2.8386 | 0.3567 | 0.2952 | 0.0684 | 0.0114 | 0.0171 | 0.0048 |
| 1995 | 2.9409 | 1.4858 | 0.2242 | 0.1918 | 0.0405 | 0.0071 | 0.0103 |
| 1996 | 2.0263 | 1.3183 | 0.6213 | 0.0947 | 0.0803 | 0.0164 | 0.0064 |
| 1997 | 3.9615 | 0.8032 | 0.5352 | 0.2414 | 0.0371 | 0.032 | 0.0083 |
| 1998 | 2.5372 | 1.7829 | 0.31 | 0.1995 | 0.0906 | 0.0149 | 0.0154 |
| 1999 | 1.8973 | 1.0928 | 0.7211 | 0.109 | 0.0728 | 0.0341 | 0.0107 |
| 2000 | 2.8374 | 0.742 | 0.4285 | 0.2624 | 0.0371 | 0.0266 | 0.0151 |
| 2001 | 1.4578 | 1.1992 | 0.2789 | 0.1555 | 0.0912 | 0.0127 | 0.013 |
| 2002 | 2.3693 | 0.5209 | 0.4738 | 0.098 | 0.0559 | 0.0332 | 0.0076 |
| 2003 | 1.2616 | 0.9991 | 0.1904 | 0.1722 | 0.0339 | 0.02 | 0.014 |
| 2004 | 1.6895 | 0.3931 | 0.3834 | 0.0649 | 0.061 | 0.0121 | 0.0104 |
| 2005 | 1.1902 | 0.582 | 0.1321 | 0.133 | 0.022 | 0.0218 | 0.0066 |
| 2006 | 1.6272 | 0.4218 | 0.2006 | 0.0412 | 0.0442 | 0.0075 | 0.0089 |
| 2007 | 0.8158 | 0.6822 | 0.1575 | 0.0661 | 0.0132 | 0.015 | 0.0045 |
| 2008 | 1.1189 | 0.3177 | 0.3179 | 0.0554 | 0.022 | 0.0044 | 0.0061 |
| 2009 | 2.0027 | 0.4693 | 0.1238 | 0.137 | 0.0226 | 0.0077 | 0.0029 |
| 2010* | 2.2547 | 0.8797 | 0.183 | 0.0444 | 0.0514 | 0.0087 | 0.0031 |
| 2011* | 3.2228 | 1.1214 | 0.3972 | 0.0692 | 0.0158 | 0.0195 | 0.0038 |
| GM(78-09) | 1.8593 | 0.6196 | 0.2262 | 0.0903 | 0.0355 | 0.0160 | 0.0092 |

*2010 and 2011 values are standard errors on TSA-derived projections of population numbers.

Table 3.2.14. Cod in Division VIa. TSA estimates for mortality-at-age.

| Year | Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 0.5165 | 0.612 | 0.6337 | 0.7597 | 0.7869 | 0.7865 | 0.782 |
| 1979 | 0.5606 | 0.7057 | 0.8686 | 1.0037 | 0.9785 | 0.9673 | 0.9523 |
| 1980 | 0.4527 | 0.6323 | 0.6798 | 0.7792 | 0.7986 | 0.7789 | 0.7717 |
| 1981 | 0.4741 | 0.6644 | 0.7537 | 0.7514 | 0.6814 | 0.7251 | 0.7355 |
| 1982 | 0.5913 | 0.6611 | 0.764 | 0.8231 | 0.8428 | 0.8368 | 0.8427 |
| 1983 | 0.673 | 0.7499 | 0.8415 | 0.9025 | 0.909 | 0.9441 | 0.9546 |
| 1984 | 0.5568 | 0.7539 | 0.8862 | 0.9578 | 1.0229 | 0.9799 | 0.9569 |
| 1985 | 0.7785 | 0.9089 | 0.9323 | 1.1425 | 1.0269 | 1.1062 | 1.0897 |
| 1986 | 0.4898 | 0.6775 | 0.8192 | 0.8925 | 0.8911 | 0.8861 | 0.8612 |
| 1987 | 0.7925 | 0.8041 | 0.9279 | 1.0716 | 1.0034 | 1.0076 | 1.0071 |
| 1988 | 0.6324 | 0.7781 | 0.9291 | 0.8839 | 0.962 | 0.9407 | 0.9265 |
| 1989 | 0.6145 | 0.7588 | 0.9695 | 1.0065 | 0.999 | 1.0216 | 1.0069 |
| 1990 | 0.5596 | 0.7336 | 0.7515 | 0.9073 | 0.834 | 0.8184 | 0.808 |
| 1991 | 0.6758 | 0.8563 | 0.8914 | 0.8765 | 0.9606 | 0.9712 | 0.9868 |
| 1992 | 0.5488 | 0.7103 | 0.9082 | 0.9892 | 0.8902 | 0.8768 | 0.898 |
| 1993 | 0.5754 | 0.7106 | 0.8945 | 0.8312 | 0.9002 | 0.8852 | 0.8785 |
| 1994 | 0.4987 | 0.6015 | 0.7262 | 0.7738 | 0.7933 | 0.7806 | 0.7951 |
| 1995 | 0.5603 | 0.7022 | 0.8104 | 0.8529 | 0.862 | 0.8637 | 0.8645 |
| 1996 | 0.6054 | 0.7412 | 0.8656 | 0.9014 | 0.8919 | 0.9059 | 0.9069 |
| 1997 | 0.593 | 0.7509 | 0.8683 | 0.9265 | 0.9177 | 0.9151 | 0.9192 |
| 1998 | 0.61 | 0.7638 | 0.8854 | 0.9348 | 0.9374 | 0.9346 | 0.9344 |
| 1999 | 0.616 | 0.7558 | 0.8799 | 0.9344 | 0.9332 | 0.9318 | 0.931 |
| 2000 | 0.5779 | 0.7245 | 0.8343 | 0.8768 | 0.8861 | 0.8883 | 0.8877 |
| 2001 | 0.622 | 0.7596 | 0.8781 | 0.9313 | 0.913 | 0.9262 | 0.927 |
| 2002 | 0.6065 | 0.7593 | 0.8836 | 0.9319 | 0.9305 | 0.9268 | 0.9302 |
| 2003 | 0.6454 | 0.7875 | 0.9013 | 0.9598 | 0.9625 | 0.957 | 0.9574 |
| 2004 | 0.6774 | 0.8039 | 0.9354 | 0.9886 | 0.9813 | 0.9836 | 0.9818 |
| 2005 | 0.6398 | 0.8204 | 0.9482 | 1.0003 | 0.998 | 0.9927 | 0.9925 |
| 2006 | 0.5303 | 0.7204 | 0.8528 | 0.9079 | 0.895 | 0.8973 | 0.8963 |
| 2007 | 0.5837 | 0.7045 | 0.7972 | 0.8793 | 0.8825 | 0.8745 | 0.8753 |
| 2008 | 0.623 | 0.7778 | 0.9177 | 0.9626 | 0.9634 | 0.961 | 0.9591 |
| 2009 | 0.5769 | 0.7583 | 0.8891 | 0.95 | 0.949 | 0.9397 | 0.9385 |
| 2010* | 0.5825 | 0.7298 | 0.8495 | 0.9088 | 0.9074 | 0.908 | 0.9059 |
| 2011* | 0.5854 | 0.7317 | 0.8488 | 0.9045 | 0.9045 | 0.9045 | 0.9045 |
| GM(78-09) | 0.5911 | 0.7362 | 0.8500 | 0.9113 | 0.9087 | 0.9094 | 0.9079 |

*Estimates for 2010 and 2011 are TSA projections.

Table 3.2.15. Cod in Division VIa. Standard errors of TSA estimates for log mortality-at-age.

| Year | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1978 | 0.194 | 0.0997 | 0.0638 | 0.0641 | 0.0768 | 0.0909 | 0.0921 |
| 1979 | 0.2021 | 0.1017 | 0.0583 | 0.0565 | 0.0681 | 0.0867 | 0.0897 |
| 1980 | 0.1983 | 0.1004 | 0.0632 | 0.064 | 0.0686 | 0.0876 | 0.0907 |
| 1981 | 0.2057 | 0.0892 | 0.0603 | 0.0627 | 0.0741 | 0.0901 | 0.0933 |
| 1982 | 0.1998 | 0.0937 | 0.0633 | 0.0652 | 0.079 | 0.0907 | 0.0979 |
| 1983 | 0.1755 | 0.0851 | 0.0601 | 0.0625 | 0.0744 | 0.0881 | 0.0928 |
| 1984 | 0.1961 | 0.0943 | 0.0617 | 0.0633 | 0.0709 | 0.0876 | 0.0927 |
| 1985 | 0.1834 | 0.0775 | 0.0629 | 0.0593 | 0.0746 | 0.0845 | 0.0906 |
| 1986 | 0.2082 | 0.0913 | 0.0635 | 0.066 | 0.0736 | 0.0926 | 0.0908 |
| 1987 | 0.1783 | 0.0911 | 0.0596 | 0.0598 | 0.078 | 0.0891 | 0.0944 |
| 1988 | 0.2053 | 0.0761 | 0.0577 | 0.065 | 0.0712 | 0.0941 | 0.0953 |
| 1989 | 0.1882 | 0.0846 | 0.0641 | 0.0609 | 0.0735 | 0.0857 | 0.0964 |
| 1990 | 0.2013 | 0.0711 | 0.0645 | 0.0657 | 0.0742 | 0.0906 | 0.0926 |
| 1991 | 0.1949 | 0.0691 | 0.0612 | 0.0636 | 0.0703 | 0.0873 | 0.0947 |
| 1992 | 0.1926 | 0.0769 | 0.064 | 0.065 | 0.0792 | 0.0881 | 0.0954 |
| 1993 | 0.205 | 0.0837 | 0.0758 | 0.0773 | 0.0867 | 0.0996 | 0.0976 |
| 1994 | 0.2159 | 0.1201 | 0.1131 | 0.1167 | 0.123 | 0.1236 | 0.1238 |
| 1995 | 0.2335 | 0.1432 | 0.1383 | 0.1385 | 0.1393 | 0.14 | 0.14 |
| 1996 | 0.2343 | 0.1432 | 0.1383 | 0.1386 | 0.1392 | 0.1399 | 0.1399 |
| 1997 | 0.2306 | 0.1456 | 0.1395 | 0.1393 | 0.14 | 0.1407 | 0.1408 |
| 1998 | 0.2341 | 0.1438 | 0.1408 | 0.1394 | 0.1401 | 0.1408 | 0.1409 |
| 1999 | 0.2349 | 0.1464 | 0.1405 | 0.1412 | 0.141 | 0.1417 | 0.1418 |
| 2000 | 0.2348 | 0.1477 | 0.143 | 0.1422 | 0.1428 | 0.1429 | 0.1429 |
| 2001 | 0.2338 | 0.145 | 0.1411 | 0.1399 | 0.1406 | 0.1413 | 0.1413 |
| 2002 | 0.2318 | 0.1465 | 0.1401 | 0.1405 | 0.1408 | 0.1414 | 0.1415 |
| 2003 | 0.2337 | 0.1441 | 0.1418 | 0.1399 | 0.1406 | 0.1413 | 0.1413 |
| 2004 | 0.2274 | 0.1462 | 0.139 | 0.1398 | 0.1402 | 0.1409 | 0.141 |
| 2005 | 0.2347 | 0.1469 | 0.1427 | 0.1408 | 0.1417 | 0.1423 | 0.1424 |
| 2006 | 0.2364 | 0.149 | 0.1438 | 0.1432 | 0.1431 | 0.1437 | 0.1438 |
| 2007 | 0.2345 | 0.1467 | 0.1428 | 0.1425 | 0.1428 | 0.1431 | 0.1432 |
| 2008 | 0.2367 | 0.1478 | 0.1406 | 0.1409 | 0.1419 | 0.1424 | 0.1425 |
| 2009 | 0.2372 | 0.1491 | 0.1438 | 0.1431 | 0.1431 | 0.1439 | 0.144 |
| 2010* | 0.2426 | 0.1525 | 0.1479 | 0.1467 | 0.1468 | 0.1467 | 0.1468 |
| 2011* | 0.2432 | 0.153 | 0.1484 | 0.1473 | 0.1473 | 0.1473 | 0.1473 |
|  |  |  |  |  |  |  |  |
| GM(78-09) | 0.2132 | 0.1114 | 0.0933 | 0.0941 | 0.1021 | 0.1122 | 0.1147 |

*Estimates for 2010 and 2011 are standard errors of TSA projections of $\log F$.
 mates for 2010, 2011 are TSA projections.

| Year | Landings (000 tonnes) |  |  | Discards (000 tonnes) |  |  | Total catch (000 tonnes) |  |  | Mean Z-0.2 (2-5) |  | SSB (000 tonnes) |  | TSB (000 tonnes) |  | Recruitment at age 1 (millions) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE |  |  |  |  |  |  |  |  |  |  |  |
| 1978 | 13.5205 | 13.2371 | 0.5744 | 3.6808 | 3.3066 | 0.8232 | 17.2013 | 17.3151 | 1.1717 | 0.6981 | 0.0313 | 26.0472 | 0.7585 | 39.5885 | 1.5141 | 20.5672 | 2.9004 |
| 1979 | 16.0887 | 15.7626 | 0.6491 | 0.0541 | 4.1523 | 0.7446 | 16.1427 | 26.7454 | 1.9934 | 0.8891 | 0.0353 | 28.5038 | 0.7836 | 56.0786 | 1.9405 | 28.3269 | 2.1581 |
| 1980 | 17.8789 | 17.2852 | 0.7873 | 0.9958 | 3.3932 | 0.8168 | 18.8747 | 23.9142 | 1.67 | 0.7225 | 0.0321 | 31.7783 | 1.0837 | 56.3863 | 1.9947 | 31.0962 | 2.5842 |
| 1981 | 23.8646 | 22.1158 | 1.3599 | 0.5198 | 0.9604 | 0.2967 | 24.3843 | 24.1205 | 1.4454 | 0.7127 | 0.0306 | 38.0061 | 1.2457 | 52.5513 | 1.8389 | 10.5034 | 1.1827 |
| 1982 | 21.5108 | 23.2089 | 1.0177 | 1.6539 | 2.3873 | 0.7303 | 23.1647 | 25.8644 | 1.5172 | 0.7727 | 0.0351 | 37.4586 | 1.1842 | 54.2507 | 1.7096 | 25.7117 | 2.1817 |
| 1983 | 21.3052 | 20.9504 | 0.9072 | 2.0195 | 1.6127 | 0.4626 | 23.3247 | 22.7346 | 1.2486 | 0.8507 | 0.0351 | 32.0595 | 1.0713 | 44.1324 | 1.5278 | 15.4779 | 1.5714 |
| 1984 | 21.2717 | 19.866 | 0.9578 | 0.6355 | 2.5622 | 0.6473 | 21.9071 | 24.3711 | 1.5578 | 0.9052 | 0.0384 | 29.9251 | 1.1271 | 48.4828 | 1.7927 | 24.0547 | 1.7581 |
| 1985 | 18.6071 | 17.6551 | 0.7976 | 8.8246 | 1.2902 | 0.3571 | 27.4317 | 17.3705 | 1.0091 | 1.0026 | 0.0401 | 22.0667 | 0.8829 | 30.2487 | 1.1673 | 12.357 | 1.4787 |
| 1986 | 11.8201 | 11.5845 | 0.6335 | 1.1998 | 1.6942 | 0.4097 | 13.0199 | 13.654 | 0.8917 | 0.8201 | 0.0361 | 18.4836 | 0.7472 | 28.8433 | 1.0713 | 19.1694 | 1.4828 |
| 1987 | 18.9705 | 18.1209 | 0.9572 | 8.7876 | 3.8839 | 1.3967 | 27.7581 | 20.9097 | 2.1193 | 0.9518 | 0.0408 | 19.6954 | 0.7284 | 39.2358 | 2.4639 | 59.8868 | 10.0221 |
| 1988 | 20.4133 | 18.5944 | 1.2724 | 1.133 | 0.8416 | 0.2962 | 21.5462 | 18.6539 | 1.3792 | 0.8883 | 0.0352 | 23.4725 | 1.0034 | 36.0528 | 1.8434 | 6.0555 | 1.1223 |
| 1989 | 17.1693 | 15.079 | 1.008 | 2.818 | 2.1561 | 0.6608 | 19.9873 | 17.1484 | 1.3328 | 0.9335 | 0.0388 | 21.0412 | 1.0995 | 32.3916 | 1.5729 | 19.7253 | 2.0397 |
| 1990 | 12.1755 | 11.9236 | 0.624 | 0.3141 | 0.3747 | 0.1339 | 12.4896 | 12.3494 | 0.7421 | 0.8066 | 0.0331 | 17.7548 | 0.711 | 24.8666 | 0.9577 | 6.3566 | 1.1508 |
| 1991 | 10.9267 | 10.7992 | 0.518 | 0.9095 | 0.8923 | 0.3199 | 11.8362 | 11.621 | 0.7487 | 0.8962 | 0.0348 | 15.2345 | 0.572 | 21.9853 | 0.9106 | 11.1191 | 1.5709 |
| 1992 | 9.0862 | 8.9169 | 0.4212 | 2.9024 | 1.3508 | 0.3727 | 11.9886 | 9.9895 | 0.6294 | 0.8745 | 0.0382 | 12.4568 | 0.4944 | 20.2179 | 0.8023 | 17.3426 | 1.645 |
| 1993 | 10.3142 | 10.421 | 0.4486 | 0.1846 | 0.7424 | 0.2303 | 10.4988 | 11.5173 | 0.6327 | 0.8341 | 0.0466 | 14.6149 | 0.623 | 23.5798 | 1.0602 | 7.1597 | 1.0106 |
| 1994 | 8.9279 | 9.1473 | 0.4347 | 0.1863 | 1.115 | 0.3715 | 9.1142 | 11.1747 | 0.8437 | 0.7237 | 0.0684 | 15.232 | 1.0902 | 25.5879 | 1.9444 | 15.0998 | 2.8386 |
| 1995 | 9.4385 | 11.0937 | 1.6807 | 0.258 | 0.893 | 0.3261 | 9.6965 | 12.5075 | 1.8966 | 0.8069 | 0.0942 | 16.8305 | 1.9405 | 26.0435 | 2.9493 | 12.6504 | 2.9409 |
| 1996 | 9.4267 | 11.6635 | 1.9608 | 0.086 | 0.451 | 0.2381 | 9.5127 | 12.5554 | 2.1394 | 0.85 | 0.0991 | 17.6804 | 2.3307 | 24.6382 | 3.298 | 4.9338 | 2.0263 |
| 1997 | 7.0336 | 9.3234 | 1.8496 | 0.3537 | 1.9584 | 0.7845 | 7.3872 | 12.1106 | 2.2333 | 0.8658 | 0.1021 | 14.1484 | 2.2096 | 24.5775 | 3.4726 | 17.6711 | 3.9615 |
| 1998 | 5.7139 | 8.9386 | 1.8062 | 0.4175 | 0.7832 | 0.3607 | 6.1314 | 9.6408 | 1.8246 | 0.8803 | 0.1037 | 11.9602 | 1.8971 | 19.0419 | 2.8589 | 8.7168 | 2.5372 |
| 1999 | 4.201 | 7.8799 | 1.6423 | 0.0879 | 0.5422 | 0.302 | 4.2889 | 8.6041 | 1.7502 | 0.8758 | 0.1043 | 11.7733 | 2.0251 | 16.6666 | 2.7623 | 4.9317 | 1.8973 |
| 2000 | 2.9771 | 6.6338 | 1.5307 | 0.6049 | 0.9827 | 0.4766 | 3.582 | 7.7634 | 1.5832 | 0.8304 | 0.1003 | 10.1766 | 1.8245 | 15.8646 | 2.5979 | 10.2059 | 2.8374 |
| 2001 | 2.347 | 6.2757 | 1.3139 | 0.2093 | 0.3406 | 0.2216 | 2.5563 | 6.4685 | 1.3211 | 0.8705 | 0.1029 | 9.0463 | 1.5097 | 12.4965 | 2.0788 | 3.249 | 1.4578 |
| 2002 | 2.2426 | 5.2256 | 1.3194 | 0.1662 | 0.8379 | 0.4228 | 2.4089 | 6.4057 | 1.3719 | 0.8763 | 0.1038 | 7.7447 | 1.4298 | 12.7508 | 2.1734 | 8.7484 | 2.3693 |
| 2003 | 1.2411 | 4.6288 | 1.0572 | 0.0458 | 0.2299 | 0.2159 | 1.2869 | 4.888 | 1.1215 | 0.9028 | 0.1066 | 6.5379 | 1.1806 | 9.277 | 1.7771 | 1.6749 | 1.2616 |
| 2004 | 0.5402 | 3.4961 | 0.9264 | 0.0718 | 0.3838 | 0.2587 | 0.612 | 3.8611 | 0.9652 | 0.9273 | 0.1089 | 5.2231 | 1.0778 | 7.1158 | 1.4919 | 3.8522 | 1.6895 |
| 2005 | 0.5114 | 2.7972 | 0.9723 | 0.0406 | 0.5699 | 0.301 | 0.552 | 3.3943 | 0.8753 | 0.9417 | 0.1124 | 3.6816 | 0.8905 | 6.5584 | 1.3334 | 5.3315 | 1.1902 |
| 2006 | 0.4545 | 2.8355 | 1.125 | 0.4777 | 0.9806 | 0.4914 | 0.9323 | 3.6305 | 0.7261 | 0.844 | 0.1023 | 3.5726 | 0.7107 | 7.9794 | 1.1192 | 11.0528 | 1.6272 |
| 2007 | 0.5242 | 4.0165 | 0.7486 | 2.0833 | 0.2933 | 0.341 | 2.6076 | 4.3412 | 0.6996 | 0.8159 | 0.098 | 5.8786 | 0.7359 | 9.0067 | 1.053 | 2.4881 | 0.8158 |
| 2008 | 0.4501 | 5.3011 | 0.9146 | 0.9084 | 0.3331 | 0.2001 | 1.3585 | 4.6258 | 0.7249 | 0.9054 | 0.1076 | 6.5852 | 0.8233 | 8.6242 | 1.0585 | 3.8602 | 1.1189 |
| 2009 | 0.222 | 5.1478 | 1.9255 | 1.3803 | 0.9711 | 0.5292 | 1.6023 | 4.5052 | 0.7821 | 0.8866 | 0.1073 | 5.1661 | 0.8043 | 9.1197 | 1.2551 | 10.39 | 2.0027 |
| 2010* | NA | 5.437 | 1.3105 |  | 0.6782 | 0.5094 |  | 5.1319 | 0.8588 | 0.8489 | 0.1061 | 6.2265 | 0.921 | 10.5334 | 1.5072 | 7.0617 | 2.2547 |
| 2011* | NA | 6.0727 | 1.4229 |  | 0.5207 | 0.433 |  | 5.6055 | 1.1157 | 0.8474 | 0.1064 | 7.6133 | 1.3545 | 11.2068 | 2.1134 | 5.5452 | 3.2228 |
| Min | 0.2220 | 2.7972 | 0.4212 | 0.0406 | 0.2299 | 0.1339 | 0.5520 | 3.3943 | 0.6294 | 0.6981 | 0.0306 | 3.5726 | 0.4944 | 6.5584 | 0.8023 | 1.6749 | 0.8158 |
| GM | 5.4059 | 9.5484 | 0.9713 | 0.5336 | 1.0035 | 0.3980 | 6.9774 | 10.8065 | 1.1878 | 0.8520 | 0.0604 | 13.8848 | 1.0511 | 21.7514 | 1.6610 | 10.2398 | 1.8593 |
| AM | 10.0367 | 11.2477 | 1.0669 | 1.3753 | 1.3521 | 0.4544 | 11.4121 | 12.9611 | 1.2796 | 0.8551 | 0.0689 | 16.8699 | 1.1436 | 26.3825 | 1.7935 | 13.7427 | 2.1385 |
| Max | 23.8646 | 23.2089 | 1.9608 | 8.8246 | 4.1523 | 1.3967 | 27.7581 | 26.7454 | 2.2333 | 1.0026 | 0.1124 | 38.0061 | 2.3307 | 56.3863 | 3.4726 | 59.8868 | 10.0221 |

Table 3.2.17. Cod in Division VIa. Inputs to short-term predictions from TSA run. Mean weights assumed from final 3 years. Note: Text is presented as it was output from WGFRANSW but data referred to as that for the human consumption fishery should be regarded as that for removals in addition to the assumed value of natural mortality.

| Label | Value | CV | Label | Value | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Number at age |  |  | Weight in the stock |  |  |
| N1 | 7061 | 0.32 | WS1 | 0.23 | 0.14 |
| N2 | 4699 | 0.19 | WS2 | 1.10 | 0.19 |
| N3 | 651 | 0.28 | WS3 | 2.42 | 0.12 |
| N4 | 144 | 0.31 | WS4 | 4.23 | 0.14 |
| N5 | 209 | 0.24 | WS5 | 5.98 | 0.06 |
| N6 | 32 | 0.27 | WS6 | 7.93 | 0.16 |
| N7 | 7 | 0.41 | WS7 | 8.64 | 0.02 |
| H.cons selectivity |  |  | Weight in the HC catch |  |  |
| sH1 | 0.59 | 0.04 | WH1 | 0.23 | 0.14 |
| sH2 | 0.75 | 0.05 | WH2 | 1.10 | 0.19 |
| sH3 | 0.87 | 0.07 | WH3 | 2.42 | 0.12 |
| sH4 | 0.93 | 0.05 | WH4 | 4.23 | 0.14 |
| sH5 | 0.93 | 0.05 | WH5 | 5.98 | 0.06 |
| sH6 | 0.93 | 0.05 | WH6 | 7.93 | 0.16 |
| sH7 | 0.92 | 0.05 | WH7 | 8.64 | 0.02 |
| Natural mortality |  |  | Proportion mature |  |  |
| M1 | 0.20 | 0.10 | MT1 | 0.00 | 0.10 |
| M2 | 0.20 | 0.10 | MT2 | 0.52 | 0.10 |
| M3 | 0.20 | 0.10 | MT3 | 0.86 | 0.10 |
| M4 | 0.20 | 0.10 | MT4 | 1.00 | 0.10 |
| M5 | 0.20 | 0.10 | MT5 | 1.00 | 0.00 |
| M6 | 0.20 | 0.10 | MT6 | 1.00 | 0.00 |
| M7 | 0.20 | 0.10 | MT7 | 1.00 | 0.00 |
| Relative effort |  |  | Year effect for natural mortality |  |  |
| in HC fishery |  |  |  |  |  |
| HF10 | 0.75 | 0.05 | K10 | 1.00 | 0.10 |
| HF11 | 0.56 | 0.05 | K11 | 1.00 | 0.10 |
| HF12 | 1.00 | 0.05 | K12 | 1.00 | 0.10 |

Recruitment in 2011 and 2012

| R11 | 4696 | 0.62 |
| :--- | :--- | :--- |
| R12 | 4696 | 0.62 |

Proportion of $F$ before spawning $=.00$
Proportion of $M$ before spawning $=.00$
Stock numbers in 2010 are TSA survivors.

Table 3.2.18. Cod in Division VIa. Results of short-term forecasts from TSA run. Management options. Note: Text is presented as it was output from WGFRANSW but data referred to as that for the human consumption fishery should be regarded as that for removals in addition to the assumed value of natural mortality.


Table 3.2.19. Cod in Division VIa. Results of short-term forecasts from TSA run. Detailed tables. Note: Text is presented as it was output from WGFRANSW but data referred to as that for the human consumption fishery should be regarded as that for removals in addition to the assumed value of natural mortality.


## Forecast for year 2011

F multiplier H.cons=0.56


Tabele 3.2.20. Cod in Division VIa. Output from srmsymc ADMB package.

| Stock name |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cod-6a |  |  |  |  |  |  |  |  |  |
| Sen filename |  |  |  |  |  |  |  |  |  |
| sum_and_sen_files/codvia10runspalyhf075hf0563.sen |  |  |  |  |  |  |  |  |  |
| pf, pm |  |  |  |  |  |  |  |  |  |
| 00 |  |  |  |  |  |  |  |  |  |
| Number of iterations |  |  |  |  |  |  |  |  |  |
| 1000 |  |  |  |  |  |  |  |  |  |
| Simulate variation in Biological parameters |  |  |  |  |  |  |  |  |  |
| TRUE |  |  |  |  |  |  |  |  |  |
| SR relationship constrained |  |  |  |  |  |  |  |  |  |
| TRUE |  |  |  |  |  |  |  |  |  |
| Ricker |  |  |  |  |  |  |  |  |  |
| 1000/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| Deterministic | 0.83 | 0.35 | 107615.00 | 33631.40 | 0.77 | 0.32 | 0.86 | $1.22 \mathrm{E}-05$ | 64.52 |
| Mean | 0.79 | 0.34 | 248654.55 | 80885.39 | 0.78 | 0.38 | 0.93 | $1.45 \mathrm{E}-05$ |  |
| 5\%ile | 0.59 | 0.26 | 42534.56 | 16130.92 | 0.61 | 0.05 | 0.68 | $1.73 \mathrm{E}-06$ |  |
| 25\%ile | 0.69 | 0.30 | 64432.03 | 23129.35 | 0.70 | 0.18 | 0.80 | 7.03E-06 |  |
| 50\%ile | 0.78 | 0.33 | 94637.85 | 32832.15 | 0.77 | 0.35 | 0.90 | $1.35 \mathrm{E}-05$ |  |
| 75\%ile | 0.88 | 0.37 | 176432.50 | 56775.68 | 0.85 | 0.53 | 1.04 | $2.02 \mathrm{E}-05$ |  |
| 95\%ile | 1.03 | 0.42 | 692590.35 | 217198.55 | 0.97 | 0.82 | 1.32 | $3.16 \mathrm{E}-05$ |  |
| CV | 0.17 | 0.15 | 3.43 | 3.41 | 0.14 | 0.65 | 0.21 | 0.65 |  |

## Table 3.2.20. (cont): Cod in Division VIa. Output from srmsymc ADMB package.

| Beverton-Holt |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1000/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| Deterministic | 0.85 | 0.18 | 401035.00 | 66296.50 | 0.39 | 1.31 | 53828.10 | 60405.70 | 64.48 |
| Mean | 0.83 | 0.17 | 830128.89 | 113018.89 | 0.54 | 1.41 | 91481.79 | 119568.27 |  |
| 5\%ile | 0.59 | 0.11 | 110359.80 | 21448.08 | 0.07 | 1.10 | 18394.14 | 11822.00 |  |
| 25\%ile | 0.70 | 0.15 | 195133.00 | 35526.05 | 0.28 | 1.26 | 28078.33 | 26150.93 |  |
| 50\%ile | 0.79 | 0.17 | 322891.50 | 55212.35 | 0.48 | 1.40 | 44006.65 | 47156.45 |  |
| 75\%ile | 0.91 | 0.19 | 630754.50 | 96558.98 | 0.76 | 1.55 | 76202.40 | 97400.13 |  |
| 95\%ile | 1.15 | 0.21 | 2769898.00 | 341061.90 | 1.15 | 1.78 | 298192.60 | 417604.45 |  |
| CV | 0.25 | 0.21 | 2.78 | 1.97 | 0.65 | 0.15 | 2.22 | 2.75 |  |

Smooth hockeystick

| 1000/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| Deterministic | 0.75 | 0.22 | 135085.00 | 27314.90 | 0.45 | 1.54 | 0.37 | 26047.10 | 64.56 |
| Mean | 0.70 | 0.21 | 173441.36 | 30090.20 | 0.47 | 1.58 | 0.38 | 26727.73 |  |
| 5\%ile | 0.53 | 0.13 | 68545.05 | 17722.69 | 0.37 | 0.99 | 0.30 | 16778.00 |  |
| 25\%ile | 0.62 | 0.19 | 98326.80 | 23808.10 | 0.42 | 1.33 | 0.34 | 22442.08 |  |
| 50\%ile | 0.69 | 0.22 | 129465.50 | 28856.20 | 0.46 | 1.58 | 0.37 | 26719.35 |  |
| 75\%ile | 0.77 | 0.24 | 171332.00 | 34618.58 | 0.50 | 1.87 | 0.41 | 31474.53 |  |
| 95\%ile | 0.89 | 0.27 | 306434.25 | 46886.99 | 0.58 | 2.17 | 0.47 | 36539.60 |  |
| CV | 0.16 | 0.22 | 1.38 | 0.31 | 0.16 | 0.23 | 0.16 | 0.23 |  |
| Per recruit |  |  |  |  |  |  |  |  |  |
|  | F35 | F40 | F01 | Fmax | Bmsypr | MSYpr | Fpa | Flim |  |
| Deterministic | 0.18 | 0.15 | 0.14 | 0.22 | 7.10 | 1.44 | 0.60 | 0.80 |  |
| Mean | 0.16 | 0.14 | 0.13 | 0.21 | 8.70 | 1.51 |  |  |  |
| 5\%ile | 0.06 | 0.05 | 0.06 | 0.13 | 3.97 | 1.07 |  |  |  |
| 25\%ile | 0.14 | 0.12 | 0.12 | 0.19 | 5.23 | 1.27 |  |  |  |
| 50\%ile | 0.17 | 0.14 | 0.14 | 0.22 | 6.48 | 1.47 |  |  |  |
| 75\%ile | 0.20 | 0.17 | 0.16 | 0.24 | 8.31 | 1.66 |  |  |  |
| 95\%ile | 0.23 | 0.19 | 0.18 | 0.27 | 15.11 | 2.16 |  |  |  |
| CV | 0.31 | 0.31 | 0.28 | 0.22 | 1.36 | 0.22 |  |  |  |



Figure 3.2.1. Cod in Division VIa. Mean weights-at-age in landings and discards. A Loess smoother has been fitted to the data at each age, with a span including three quarters of the data points.

## Catch



Figure 3.2.2. Cod in Division VIa. Log catch (landings + discards) curve gradient plot using WG commercial catch-at-age data. Solid line shows time-series of gradient of linear fit to curve over the age range 2-5, dashed line over the ages 2-4 and dotted line over the ages 3-5. An increasing value indicates increasing mortality.


Figure 3.2.3. Cod in Division VIa. Cpue numbers for fish aged at 1+ by ICES statistical rectangle resulting from quarter four surveys. Scottish quarter four ground fish survey (ScoGFS-4Q) and Irish ground fish survey (IGFS). Numbers are standardised to 30 minutes towing. Irish Survey values are for fish $>23 \mathrm{~cm}$ in length (proxy for age 1+).


Figure 3.2.3. cont. Cod in Division VIa. Cpue numbers for fish aged at 1+ by ICES statistical rectangle resulting from Scottish quarter one survey (ScoGFS-1Q). Numbers are standardised to 30 minutes towing.


Figure 3.2.4. Cod in Division VIa. Log catch curves from Scottish quarter one ground fish survey (ScoGFS-1Q); ages 1-6.

## ScoGFSQ1



Figure 3.2.5. Cod in Division VIa. Log catch curve gradient plot using ScoGFS-1Q index data. Solid line shows time series of gradient of linear fit to curve over the age range 2-5, dashed line over the ages 2-4 and dotted line over the ages $3-5$. An increasing value indicates increasing mortality.


Figure 3.2.6. Cod in Division VIa. Summary plot of TSA update run. (landings \& discard data excluded from 1995 onward). Solid line in top left frame indicates removals resulting from mortality over and above $M=0.2$; open circles represent reported catch. Solid line in top right frame indicates mortality over and above $\mathrm{M}=0.2$. Dashed lines show $\pm 2$ s.e. (approx $95 \%$ confidence interval).


Figure 3.2.7. Cod in Division VIa. TSA final run. Standardised prediction errors at age plots for landings.


Figure 3.2.8. Cod in Division VIa. TSA final run. Standardised prediction errors at age plots for discards.


Figure 3.2.9. Cod in Division VIa. TSA run. Standardised prediction errors at age plots for ScoGFS-1Q.




Figure 3.2.10. Cod in Division VIa. Retrospective plots of TSA run. Biological reference points are given by horizontal dashed lines. Confidence intervals for the run using all years of data are shown by dotted lines.


Figure 3.2.11. Cod in Division VIa. TSA final run. Stock-recruit relationship. Numbers indicate year class.


Figure 3.2.12. Cod in Division VIa. Precautionary approach plot. Mortality is all mortality over and above the fixed natural mortality value of 0.2 (referred to as ' $\mathrm{Z}-0 . \mathbf{2}^{\prime}$ ).

Figure Cod, Viz Short term forecast


Figure 3.2.13. Cod in Division VIa. Short-term forecast. Figure shows mortality from all sources that is over and above $M=0.2$ and associated removals.


Figure 3.2.14. Cod in Division VIa. Sensitivity analysis of short-term forecast. Removals are associated with mortality from all sources over and above $\mathbf{M}=\mathbf{0} .2$.


Deffonlecimew

Figure 3.2.15. Cod in Division VIa. Probability profiles for short-term forecast. Removals are associated with mortality from all sources over and above $\mathbf{M}=\mathbf{0 . 2}$.


Figure 3.2.16. Cod in Division VIa. Ratio of estimated to observed catch using TSA. Bars show $\pm 2$ s.e. TSA excludes catch data from 1995 to 2008 inclusive. The 'catch' resulting from TSA is considered removals from both fishing and natural mortality over and above $M=0.2$.


Figure 3.2.17. Cod in Division VIa. Stock-recruit relationships fitted by srmsymc package. Models were fitted using 1000 MCMC re-samples. Left-hand panels illustrate confidence intervals. Righthand panels present curves plotted from the first 100 re-samples for illustration. The blue line indicates a deterministic estimate, separate from the MCMC chain. The legends for each recruitment model show it was possible to converge on a value of FMSY and Fcrash for all 1000 iterations in each case.

## Cod-6a Ricker



Figure 3.2.18. Cod in Division VIa. srmsymc package. Estimation of $F$ reference points and equilibrium yield and SSB against mortality using Ricker recruitment model. For yield and SSB plots left-hand panels illustrate confidence intervals. Right-hand panels present curves plotted from the first 100 re-samples for illustration. The blue line indicates a deterministic estimate, separate from the MCMC chain. Circles show data points with the most recent year labelled. For VIa cod the model has been run using total removals over and above natural mortality, i.e. the x-axis represents Z-0.2.


Figure 3.2.19. Cod in Division VIa. srmsymc package. Estimation of $F$ reference points and equilibrium yield and SSB against mortality using Beverton-Holt recruitment model. For yield and SSB plots left-hand panels illustrate confidence intervals. Right-hand panels present curves plotted from the first 100 re-samples for illustration. The blue line indicates a deterministic estimate, separate from the MCMC chain. Circles show data points with the most recent year labelled. For VIa cod the model has been run using total removals over and above natural mortality, i.e. the $x$-axis represents $Z$ 0.2.


Figure 3.2.20. Cod in Division VIa. srmsymc package. Estimation of $F$ reference points and equilibrium yield and SSB against mortality using smooth hockey stick recruitment model. For yield and SSB plots left-hand panels illustrate confidence intervals. Right-hand panels present curves plotted from the first 100 re-samples for illustration. The blue line indicates a deterministic estimate, separate from the MCMC chain. Circles show data points with the most recent year labelled. For VIa cod the model has been run using total removals over and above natural mortality, i.e. the x-axis represents $\mathrm{Z}-0.2$.


Figure 3.2.21. Cod in Division VIa. srmsymc package. F reference points and yield-per-recruit and SSB-per-recruit against mortality. For VIa cod the model has been run using total removals over and above natural mortality, i.e. the $x$-axis represents Z-0.2.


Figure 3.2.22. Cod in Division VIa. Comparison of SSB, recruitment-at-age one and mean F (2-5) estimates produced by final run assessments between this year's assessment and assessments going back to 2001.


Figure 3.2.23. Cod in Division VIa. Summary plot of TSA run where landings \& discard data excluded from 1995-2005. Solid line in top left frame indicates removals resulting from mortality over and above $\mathrm{M}=0.2$; open circles represent reported catch. Solid line in top right frame indicates mortality over and above $\mathrm{M}=0.2$. Dashed lines show $\pm 2$ s.e. (approx $95 \%$ confidence interval).


Figure 3.2.24. Scottish (and other EU landing into Scotland) VMS data on fishing activity using TR1 gear (VMS pings per square n.m.) associated with cod landings from daily logbook entries (colour scale). Overlaid are ScoGFS-1Q survey cpues centred on the statistical rectangle sampled. Dashed lines show ICES divisions, the broken line represents the cod management line and the solid line shows the limits of the UK EEZ, highlighting the extent of EU waters in Subdivision Vb. Depth contours are at 200 m intervals.


Figure 3.2.25. Scottish (and other EU landing into Scotland) cod landings from daily logbook entries allocated to VMS fishing pings (colour scale). Vessels using TR1 gear. Overlaid are ScoGFS1Q survey cpues centred on the statistical rectangle sampled. Dashed lines show ICES divisions, the broken line represents the cod management line and the solid line shows the limits of the UK EEZ, highlighting the extent of EU waters in Subdivision Vb. Depth contours are at 200 m intervals.

### 3.3 Haddock in Division VIa

## Type of assessment in 2010

The stock assessment of VIa haddock in 2010 is an update of last year's assessment with the TSA model, using catch data up to 1994 and tuning data from two Scottish groundfish surveys. In this year's assessment catch data were also included for the period 2006-2009 as these were thought to be recent years where sufficiently reliable catch data were available. See Section 3.3.2 for further explanation.

## ICES advice applicable to 2009

The form of ICES' advice changed in 2003 to take more account of the mixed nature of the fisheries exploiting haddock. Management of haddock since then has been considered as part of wider concerns in the Celtic Sea and West of Scotland ecosystem.

The advice relating to the single-stock exploitation boundary for 2009 was:
"Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects:

The current fishing mortality (2007) is estimated to be 0.56, which is above the rate expected to lead to high long-term yields and low risk of stock depletion.

Exploitation boundaries in relation to precautionary limits
Even in the absence of fishing the stock is not expected to be rebuilt to Bpa."

## ICES advice applicable to 2010

The advice relating to the single-species exploitation boundary for 2009 was:
"Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects:

The current fishing mortality (2008) is estimated to be 0.46, which is above the rate expected to lead to high long-term yields and low risk of stock depletion.

Exploitation boundaries in relation to precautionary limits:
In the absence of fishing, the stock is expected to be rebuilt close to $B_{p a}$ in the short term."
Following a request to evaluate a management plan for haddock in VIa:
"ICES advises that a harvest rule with a target fishing mortality of 0.3 and a TAC constraint of $\pm 15 \%$ is consistent with the precautionary approach (high probability of SSB being above Blim by 2015 and beyond). In addition, simulations suggest that this harvest rule has the best chance, among those tested, of producing a combination of low risk to biomass and high cumulative yield, thus it conforms with the goal of achieving long-term maximum sustainable yield from the stock.

The harvest rule was tested for several combinations of target fishing mortality ( $0.2,0.3,0.4$ ) and interannual variation in TAC $( \pm 15 \%, \pm 20 \%, \pm 25 \%)$. ."

Note that the statement above refers to a management plan where, when SSB is below Blim, the fishing mortality should be 0.1. Subsequent evaluations are being carried out for a management plan where the TAC constraint is $\pm 25 \%$ whether above or below Blim.

### 3.3.1 General

## Stock description and management units

A TAC relating to this stock is in place for EU and international waters of ICES management Areas Vb and VIa and the assessment is carried out using data from VIa. The basis for the stock assessment area is described in the Stock Annex.

The agreed minimum landing size for haddock in Division VIa is 30 cm . There is no formal management plan currently in place. Further regulations implemented for the west of Scotland, including technical measures associated with the cod recovery plan and the UK Registration of Buyers and Sellers regulation, are described in the overview section for this management area (Section 3.1).

The following table summarises EC TACs applied for haddock in Division VIa during 2009.

| Species: Haddock <br> Melanogrammus acglefinus | Zone:EC waters of Vb and Vla <br> (HAD/5BC6A.) |  |  |
| :--- | ---: | :---: | :---: |
| Belgium | 4 |  |  |
| Germany | 5 |  |  |
| France | 194 | 576 |  |
| Ireland | 2737 |  |  |
| United Kingdom | 3516 |  |  |
| EC | 3516 | Analytical TAC <br> TAC | Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

## Values are tonnes.

The following table summarises EC TACs applied for haddock in Division VIa during 2010.

| Species:Haddock <br> Melanogrammus aeglefinus | Zone:EU and international waters of Vb and Vla <br> (HAD $/ 5 \mathrm{BC} 6 \mathrm{~A})$ |  |
| :--- | :--- | :--- | :--- |
| Belgium | 3 |  |
| Germany | 4 |  |
| France | 147 |  |
| Ireland | 438 |  |
| United Kingdom | 2081 |  |
| EU | 2673 | Analytical TAC |
| TAC | 2673 |  |

[^1]
## Fishery in 2009

Official (reported) landings for each country participating in the fishery are given in Table 3.3.1. Vessels operating in the fishery are mainly Scottish and Irish and the amount of quota allocated to different countries reflects this.

Uptake of quota is given here and is calculated from the official landings as a proportion of the EC allocated quota for each country. No country took up its full allocated quota in 2009, although uptake was higher than in 2008. Uptake of quota has generally been low in recent years (e.g. $\sim 73 \%$ in 2006; $\sim 51 \%$ in $2007 ; \sim 45 \%$ in 2008 and $\sim 79 \%$ in 2009). Discards data that are reported are dealt with in the following section.

| COUNTRY | TAC 2009 | OFFICIAL LANDINGS* | \% UPTAKE OF QUOTA |
| :--- | :---: | :---: | :---: |
| Belgium | 4 | 0 | $0 \%$ |
| Germany | 5 | 0 | $0 \%$ |
| France | 194 | 124 | $64 \%$ |
| Ireland | 576 | 297 | $52 \%$ |
| UK | 2737 | 2361 | $86 \%$ |
| EC | 3516 | 2782 | $79 \%$ |

Values of TAC (Total Allowable Catch) and landings are in tonnes.

* The official landings provided to the WG for 2009 are preliminary at time of writing in 2010.


### 3.3.2 Data

An overview of the data that have been provided to the WG is given in Section 2, including sampling levels by country for this stock. The reliability of catch data for this stock was a concern for several years, due to issues such as misreporting or underreporting and associated unaccounted discarding. It became impossible to quantify the extent of unallocated removals, leading to the use at the 2006 meeting of a modified TSA assessment method which did not use catch data after 1994.

Recent changes in regulations and fleet behaviour have improved the quality of catch data, which is now thought to be more representative of the true catch. The UK Registration of Buyers and Sellers Regulations are likely to have reduced or largely eliminated underreported landings. Information from the Compliance section of Marine Scotland suggests that approximately 60 tonnes of haddock were misreported out of Area VIa in 2009 ( $<3 \%$ of the officially reported UK(Scotland) landings). There has been a significant reduction in effort in Division VIa (STECF 2009) and the TAC in recent years does not appear to be restrictive for this fishery, diminishing the incentive to underreport landings.

## Landings

Official landings as reported to ICES and estimated by the WG are provided in Table 3.3.1.

## Catch-at-age data

Total catch-at-age data (landings and discards) are given in Table 3.3.2., while catch-at-age data and mean weights-at-age for each catch component (landings and discards) are given in Tables 3.3.3-3.3.7. The full available year and age range are given for completeness: however, it should be noted that commercial catch data before 1978 are not used in the assessment, as the split of total catch into landings and discards was hypothetical prior to that time. The countries that provide data, including sam-
pling levels achieved, are listed in Table 2.1.

## Discards

WG estimates of discards are based on data collected in the Scottish and Irish discard programmes; raised by weighted average to the level of the total international discards (Table 3.3.4.). Discards data from Scotland were raised from 17 sampled trips in 2009, spread evenly across each quarter. Irish discards data were raised from three trips: adequate coverage given the lower amount of effort by Irish vessels.

## Biological

## Weights-at-age

The estimated weights-at-age for the total catch in Division VIa are given in Table 3.3.5. These are calculated as weighted averages of the corresponding weights-at-age in landings and discards: the latter are given in Tables 3.3.6.and 3.3.7. Weights-at-age in the stock are assumed to be equal to the weights-at-age in the total catch, in the absence of a sufficiently long time-series of survey-based weight measurements. The weights-at-age time-series are also plotted in Figures 3.3.1-3.3.3. These show that weights-at-age in landings (and, by extension, catch and stock) for fish aged 3 and older have declined considerably over the last $\sim 20$ years. The weights-at-age of younger fish have increased in 2009. Weights-at-age in discards are relatively constant. The supplied data for fish weights-at-age 1 in 2009 in Irish landings was 460 g . This was thought to be high in comparison to fish at age 1 in recent years and also with fish at age 2 in 2009. This datum was replace with a mean of the most recent three years of Irish fish weights at age 1, to give an estimate of 344 g . It was noted that haddock are likely to grow faster in the southern area of Division VIa, where Irish fishing vessels are most likely to operate.

## Natural mortality and maturity

Natural mortality was assumed to be 0.2 for all ages and years, and maturity was assumed to be as follows:

|  | AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3 +}$ |
| :--- | :---: | :---: | :---: | :---: |
| Proportion mature | 0.00 | 0.57 | 1.0 |  |

Proportions of $F$ and $M$ before spawning were both set to 0.0 , in order to generate abundance (and hence SSB) estimates dated to January 1st.

## Surveys

## Research vessel surveys

Four research-vessel survey series are available for the assessment of haddock in Division VIa as given in the following table:

| SURVEY | Years available | AGES AVAILABLE | AGES USED |
| :--- | :---: | :---: | :---: |
| Scottish groundfish Q1 | $1985-2010$ | $1-8$ | $1-7$ |
| Scottish groundfish Q4 | $1996-2009$ | $0-7$ | $1-7$ |
| Irish groundfish survey | $1993-2002$ | $0-8$ | - |
| New Irish groundfish survey | $2003-2009$ | $0-10$ | - |

The reports of the 2006 meeting of the WG (WGNSDS 2006) and the 2007 meeting of the IBTS WG (IBTSWG 2007) explored available survey data in detail. Both ScoGFS

Q1 and Q4 were first accepted for use in the 2006 assessment, and this practice has been continued in subsequent years. The IreGFS series was not considered further due to problems with internal consistency (ICES-WGNSDS 2006). The new IRGFS series has seven years of data and can be considered for tuning purposes at the next benchmark assessment.

All survey series available for tuning the assessment are given in Table 3.3.8, with the data that were used in the final assessment indicated in bold type. Plots of the spatial distribution of the ScoGFS Q1 and Q4 survey mean catch rates per ICES statistical rectangle by age class are given in the Stock Annex.

## Commercial cpue

## Commercial catch-effort series

The available commercial effort and lpue data for this stock are indicated in the Stock Annex.

### 3.3.3 Historical stock development

The model used for this assessment is the state space model TSA, with data from two research vessel surveys (1978-2010) and with catch data included 1978-1994 and 2006-2009, corresponding to the time periods when catch data are thought to be reliable. The model is run using a custom made Fortran 90 programme (see Stock Annex). Outputs from the TSA assessment are shown in Figures 3.3.4-3.3.10 and Tables 3.3.10-3.3.14.

The reliability of catch data for haddock was a concern for several years, and since it was not possible to quantify the extent of unallocated removals, this lead, at the 2006 meeting, to the use of a modified TSA assessment method which did not use catch data after 1994. This remained the accepted assessment method for the 2007-2009 meetings. In 2010, measurable improvements in the reliability of catch data (Section 3.3.2) led the WG to question the continued discrepancy between the prediction of landings by the model and the reported catches after 2005. Furthermore, while the assessment was primarily survey based, the uncertainty around estimates of F was appreciable, and the estimate was not coming down in years when evidence of reduced effort indicated a probable reduction in F.

The re-inclusion of catch data has been implemented with TSA in previous assessments for which this model is used. For example, catch data were re-included in the assessment of VIa cod at the 1997 meeting of the Working Group for the Assessment of Northern Shelf Demersal Stocks (WGNSDS, 1997). The catch data for cod were reincluded in following assessments, but were removed again subsequently because of more recent concerns over reported landings for that stock. See Section 3.2.

## Final update assessment

The assessment in 2010 was an update, including data indicated in the table below, which summarises the data ranges used in recent assessments.

| Data | $2006$ <br> ASSESSMENT | $2007$ <br> ASSESSMENT | $2008$ <br> ASSESSMENT | $2009$ <br> ASSESSMENT | $2010$ <br> ASSESSMENT |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Catch data | $\begin{aligned} & \text { Years: 1978- } \\ & 1994 \end{aligned}$ | $\begin{aligned} & \text { Years: 1978- } \\ & 1994 \end{aligned}$ | $\begin{aligned} & \text { Years: 1978- } \\ & 1994 \end{aligned}$ | $\begin{aligned} & \text { Years: 1978- } \\ & 1994 \end{aligned}$ | Years: 1978-1994 <br> and 2006-2009 |
|  | Ages: 1-8+ | Ages: 1-8+ | Ages: 1-8+ | Ages: 1-8+ | Ages: 1-8+ |
| Survey: <br> ScoGFS Q1 | $\begin{aligned} & \text { Years: 1985- } \\ & 2006 \end{aligned}$ | $\begin{aligned} & \text { Years: 1985- } \\ & 2007 \end{aligned}$ | $\begin{aligned} & \text { Years: } 1985- \\ & 2008 \end{aligned}$ | $\begin{aligned} & \text { Years: 1985- } \\ & 2009 \end{aligned}$ | Years: 1985-2010 <br> Ages 1-7 |
|  | Ages: 1-7 | Ages 1-7 | Ages 1-7 | Ages 1-7 |  |
| Survey: <br> ScoGFS Q4 | $\begin{aligned} & \text { Years: 1996- } \\ & 2005 \end{aligned}$ | $\begin{aligned} & \text { Years: 1996- } \\ & 2006 \end{aligned}$ | $\begin{aligned} & \text { Years: 1996- } \\ & 2007 \end{aligned}$ | $\begin{aligned} & \text { Years: 1996- } \\ & 2008 \end{aligned}$ | Years: 1996-2009 <br> Ages 1-7 |
|  | Ages: 1-5 | Ages 1-7 | Ages 1-7 | Ages 1-7 |  |
| Survey: IGFS | Not used | Not used | Not used | Not used | Not used |

Table 3.3.9 shows the evolution of the corresponding TSA parameter estimates since 2003.

Standardised prediction errors from the assessment model are shown in Figures 3.3.5 (landings), 3.3.6 (discards), 3.3.7 (ScoGFS Q1) and 3.3.8 (ScoGFS Q4). TSA is a state space model, and these prediction errors are an analogous (but not completely equivalent) diagnostic tool to residuals of fits from other stock assessment models. The small, negative prediction errors for the landings and discards in the period 2006-2009 at various ages show that the model is predicting landings and discards to be slightly higher than observed data. Generally the prediction errors do not show a pattern persisting for longer than 5 years. The only cases where this occurs are for age 1 of the ScoGFS Q1 index (Figure 3.3.7) where in 2010 the pattern is beginning to reverse. The magnitude of these (age 1 ScoGFS) prediction errors is relatively small (ranging from -0.9 to -1.6 ). A similar, inconsequential, pattern is seen in the fit to the ScoGFS Q4 index (Figure 3.3.8). None of the prediction errors are of a magnitude or show a pattern which would invalidate the model fit. Negative prediction errors in the survey indices at age 1 indicate lower than expected recruitments in recent years.

Previous assessments have applied a down-weighting to certain data points, based on the TSA prediction errors. These are described here. A notable prediction error occurred in the ScoGFS Q4 in 2007 at age 2 (the 2005 year class). This was due to a large index value in this survey year at that age (which was common to many hauls): the model setting, $q_{\text {catch }}($ age $=2$, year $=2007$ ) was altered (multiplied by 9.0 in the appropriate model settings) in order to decrease the influence of this extreme value (an adjustment recommended in Fryer 2001 which has been applied previously to several age/year data points). A prediction error from the ScoGFS Q1 in 2009 (age 4) was also down weighted according to the same procedure. No further down-weighting was applied in 2010.

There is a poor relationship between stock size (SSB) and recruitment for this stock, with large values for recruitment possible at small stock sizes and small recruitments possible at large stock sizes (Figure 3.3.9). The TSA stock-recruit plot is shown in Figure 3.3.9.

Estimated and observed discard rates (proportions-at-age) are shown in Figure 3.3.10. The discard model fits are good for the years when catch data are included (19781994 and 2006-2009) and also most other years. The observed proportions deviate slightly in 2003-2005.

TSA estimates a discard ogive for every year. However, when there are no catch data, the estimated ogive will simply be some weighted average of the discard ogives in
neighbouring years. So, when several years of catch data are omitted, the estimated discard ogives in this period will hardly change at all because there are no new data included from which to produce a new estimate. From 2006, when the catch data are re-included, the model is able to much better estimate the discard ogive (Figure 3.3.10).

## Retrospective analysis

Most retrospective bias in this stock assessment (see Figure 3.3.11) is thought to be caused by mismatch between catch and survey data (WGMG 2007), and as only survey data are used in the TSA model between 1995 and 2005 the retrospective pattern observed in F is not surprising. There is also a pattern in SSB estimates during the 1990s, and this is likely to correspond to a period when neither survey used in the assessment was able to track year-class strength well.

## Comparison with previous year's assessment

The 2009 VIa haddock assessment estimated F in 2008 at 0.46 and SSB (January 1st 2008) at 30436 tonnes. The current assessment has revised these figures, to a fishing mortality of 0.38 in 2008 and an SSB (January 2008) as 22114 tonnes ( $27 \%$ decrease). Recruitment in 2008 has been revised from 6.6 million to 7.8 million ( $\sim 18 \%$ increase).

The estimate of SSB in January 2009 from this assessment is 16818 tonnes with a standard error of 1615 tonnes ( $\sim 10 \%$ ). Last year's assessment put this figure at 20271 tonnes.

The current assessment's estimate of SSB (for January 2010) is 13337 tonnes. The short-term forecast from last year's assessment predicted SSB in 2010 to be at 15400 tonnes. This is a difference of 2063 tonnes ( $\sim 13 \%$ decrease in the estimate).

## State of the stock

The state of the stock is summarised in Figure 3.3.4 and Table 3.3.14.
The final estimates for the stock in 2009 are:

$$
\begin{array}{ll}
\mathrm{F}_{(2-6)} & =0.30 \\
\mathrm{SSB} & =16818 \mathrm{t}
\end{array}
$$

Based on the most recent estimates of SSB in 2010 (13 336 tonnes, <Blim) ICES classifies the stock as being at risk of reduced reproductive capacity.

Based on the most recent estimate of fishing mortality in $2009\left(0.30,<\mathrm{F}_{\mathrm{pa}}\right)$ ICES classifies the stock as being harvested sustainably.

Summaries from the final assessment, including, total removals, landings, discards, recruitment, mean F and SSB are given in Figure 3.3.4, while corresponding estimates and standard errors are presented in Tables 3.3.10 and 3.3.11 (population abundance), Tables 3.3.12 and 3.3.13 (fishing mortality), and Table 3.3.14 (stock summary). Mean $F_{2-6}$ is estimated to have risen to just above $F_{\mathrm{pa}}(0.5)$ during 2003-2007, subsequently falling below 0.5 in 2008, and remaining below $\mathrm{F}_{\mathrm{pa}}$ in 2009. A sequence of low recruitments led to a fall in SSB from the peak in 2003. The assessment estimates that SSB has been below $\mathrm{B}_{\mathrm{pa}}$ since 2005. The most recent estimate of recruitment, from the 2010 Quarter 1 Scottish Groundfish survey (the 2009 year class) is higher than in the last three years, but is probably below the long-term average

Uncertainty in fitted and observed catches increases from 1995-2005 (Figure 3.3.4),
which is the period when the landings and discards are excluded from the model and the survey data is used for estimation. Catch data tend to have more precision than survey data and although both survey used in the assessment have been seen to track year-class strength well, the survey data are more "noisy" (show greater variability) than the catch data. Therefore, when the catch is included in the later part of the timeseries (2006-2009) the confidence intervals of the estimates are seen to reduce.

The difference between observed and predicted catch represents unaccounted removals, amounting to about $10 \%$ of the landings by 2006-2009. The reported catch in 2009 is within the bounds of error of the estimated catch. This is thought to reflect beneficial effects of management regulations and changes in fleet behaviour since 2006, and is supported by anecdotal information from the fishing industry. For example, there has been great effort reduction by the whitefish otter trawler fleet in Di vision VIa and the TAC does not appear to be restrictive for this fishery, diminishing the incentive to under-report landings. Information from the Compliance section of Marine Scotland put estimates of area misreporting out of VIa at approximately 60 tonnes in 2009. Given that total landings were 2800 tonnes for 2009 this represents $2 \%$ of the estimate of unallocated removals. The remaining $8 \%$ of these unallocated removals ( $\sim 224$ tonnes), still within the bounds of error of the assessment model estimate, could represent (in unknown proportions) uncertainty in the estimate of misreporting and associated unreported discards, or natural mortality not accounted for by the assessment model's assumed value of 0.2.

### 3.3.4 Short-term projections

## Recruitment estimates

The TSA assessment model provides estimates of recruitment for the forecast years 2010 and 2011. The value for 2010 (that is, the 2009 year class at age 1 ) is based largely on the ScoGFS Q1 datum for 2010 (along with a degree of time-series smoothing), and as it is based on observations it is appropriate to use it in the forecast. The value for 2011 (that is, the 2010 year class at age 1) is not generated directly by data, but rather the underlying Ricker stock-recruit model that is included by TSA (Figure 3.3.9) as part of the overall model fit. As with the assessment of last year, a long-term (19782009) geometric mean is used for subsequent years (2012). The recruitment values used in the forecast are given in the following table:

| YEAR | TSA | GM (78-09) |
| :---: | :---: | :---: |
| 2010 | $41994(\sim$ ScoGFS $)$ |  |
| 2011 | 76211 (Ricker) |  |
| 2012 | - | 75905 |

There is close agreement between the TSA-generated recruitment estimates, and the indices from the two surveys (see Figure 3.3.12).

TSA produces short-term forecasts as part of every standard model run. The model will also forecast fishing mortality rates. It does so by iterating forward the timeseries model that had been fitted to historical data. These forecast mortalities therefore retain the time-series characteristics of the preceding data. Although the TSA estimates are likely to follow a pattern of damped oscillation towards an eventual steady state, the WG preferred to use standard tools (i.e. MFDP) as the basis for the forecast. The procedure used instead of TSA's built in procedure is described below.

The time-series at age of fishing mortality estimate is shown in Figure 3.3.13, along with the mean F over ages $2-6$. As with last year's assessment, a three-year mean fish-ing-mortality selection pattern was used in the forecast. Figure 3.3.14 compares a simple three-year mean, the most recent estimate (2009), and TSA-generated selection patterns.

The forecasts presented in this Section have been given as forecasts of total removals, split subsequently into removals due to landings, discards and unallocated removals (other than those assumed to be due to current estimates of natural mortality) respectively. As highlighted previously, the assessment is survey-based from 1995 to 2005 and can only estimate total removals during this period. The difference between reported and estimated catches represents unallocated removals, reflecting our uncertainty in natural mortality and a certain amount of likely area-misreporting. In the period when the assessment is survey based only the estimated amount of unallocated removals is appreciable. The 1999 year class of haddock was strong, and survey estimates of that year class would have contributed to high model estimates of predicted catch between 2002 and 2005 (Figure 3.3.4).

The WG considered that the most appropriate level of discarding to use in the forecast was a mean of the last three years. It is not possible to know what discarding practices will be in the immediate future, although since the incoming 2009 year class has been estimated to be at appreciable numbers by the Scottish and Irish groundfish surveys in Q4 2009 and by the Scottish groundfish survey in Q1 2010, it is likely that some amount of discarding will occur. The is no strong trend in discard behaviour in the last three years so taking a 3-year mean is the most unbiased approach. For the short-term forecast, the assumption is that this input $F$ remains constant.

The final key issue for the forecast is that of weights-at-age, and in particular, the slow growth observed in recent year classes. Figure 3.3.15 demonstrates this with linear models fitted to cohort-based mean weights-at-age data. A number of recent year classes appear to be growing more slowly than has been the case in the more distant past. As with last year, linear models were used as the basis for predictions for those cohorts with sufficient data (Table 3.3.15), with the small change that the models were fit using data from age $0-8+$, as this slightly improved precision (Jaworski, WD12).

Table 3.3.16 presents the inputs to the short-term forecast. Outputs from the forecast are given in Tables 3.3 .17 (management options) and Figures 3.3.16 (sensitivity analysis), 3.3.17 (probability profiles) and 3.3.18 (short-term forecast). Figure 3.3.16 shows the sensitivity of the forecast to the various input parameters; indicating that numbers at age 1 in 2010 and recruitment in 2011 are responsible for a large proportion of the variance associated with the forecast. Figure 3.3 .17 shows probability profiles for the forecast, indicating the probability of F being greater than F (status quo) in 2011 with increasing rates of total removals (left figure) and the probability of SSB being below reference biomass (e.g. reference points, $\mathrm{B}_{\mathrm{pa}}$, Blim) in 2011. This figure indicates that there is over $90 \%$ probability that SSB will be below $\mathrm{B}_{\mathrm{pa}}$ in 2012. Figure 3.3.18 shows a summary of the forecast results, indicating how SSB is projected to change with increasing rate of removals, under fishing mortality between $0-0.45$.

Results of the forecast at status quo $F$ are summarised in the following table:

| YEAR | Removals (000 T) | SSB (000 T) |
| :---: | :---: | :---: |
| 2010 | 5.35 | 13.7 |
| 2011 | 6.06 | 13.6 |
| 2012 | - | 19.9 |

At the status quo rate of removals, and given assumptions about growth and recruitment, the most recent estimate of SSB (2010) is below $\mathrm{Blim}_{\mathrm{lim}}$ and is forecast to increase in 2011 and 2012, primarily due to the most recent estimate of recruitment in 2010 being relatively high compared to those of four out of five of the most recent years.

### 3.3.5 MSY evaluations

ICES changed the basis of its advice in 2010, with estimates of $\mathrm{F}_{\text {msy }}$ being introduced. A package developed for the purpose (srmsymc) was run, using the current year's .sum and .sen files as inputs. The method used is described in Section 2.2. The current assessment puts FMSY in the range, 0.19-0.35 (Figures 3.3.19-3.3.22 and Table 3.3.18). These estimates came from fits of the Beverton-Holt (0.19), Ricker (0.31) and Smooth Hockey Stick ( 0.35 ) stock-recruit models. There were high coefficients of variation for the three stock-recruitment models used as the basis for the analysis. Furthermore, the Akaike information criterion (AIC), often used to determine statistical model fit, was almost the same for each stock-recruit model. This reflects the poor relationship between stock and recruitment, allowing the three models to be fit to stock and recruit data with comparable uncertainty. For each stock-recruit model a high (>950 out of 1000) number of model simulations were accepted by the package.

### 3.3.6 Biological reference points

ICES has defined the following reference points for this stock.

| Reference point | TeChnical basis |
| :--- | :--- |
| Bpa $=30,000 \mathrm{t}$ | Blim $^{*} 1.4$ |
| Blim $=22,000 \mathrm{t}$ | Lowest observed SSB when reference point was establised (1998) |
| Fpa $=0.5$ | High probablity of avoiding SSB falling below Bpa in the long term |
| Flim | Not defined |

### 3.3.7 Management Plans

There is no agreed management plan currently in place for this stock. ICES has evaluated a proposed management plan, the details of which can be found at:
http://www.ices.dk/committe/acom/comwork/report/2010/Special Requests/EC haddock management

### 3.3.8 Uncertainties and bias in assessment and forecast

## Quality of the assessment

## Landings and discards

Quotas for haddock in Division VIa appear to have started to become restrictive in or around 1995. Anecdotal evidence suggests that these and other strict management measures led to increasing unreliability of landings data from the commercial fleets prosecuting the fishery from 1995 to 2005. The approach taken by this WG from 2006 onwards was to assess the stock using a modified TSA model which did not include catch data from 1995 onwards, and which thus modelled removals rather than catches. During the period when the catch is not included (1994-2005) the discard ogives estimated by the model are weighted averages of those of neighbouring years. This results in little change in the estimated discard ogive in the years when the catch is excluded and an observable discrepancy between the model's discard ogive and
the reported discards proportions in 2003-2005.

## Effort

Currently commercial cpue or lpue data cannot be used in the assessment with any confidence. The assessment is therefore primarily survey-based, with landings and discards data used prior to 1995.

## Surveys

A survey-based assessment can only be as good as the surveys on which it is based. The Scottish groundfish survey series appear to have good internal consistency and to track cohorts well, with the exception of a period during the mid-1990s. Concerns remain over the apparent differences in catchability of young fish between the Scottish and Irish components of IBTS (ICES-IBTSWG 2007). Any survey is likely to become less reliable when stock abundance declines, and this issue needs to be revisited in the near future for haddock and many other stocks.

This assessment is survey based for the years 1995-2005. Re-including catch data for 2006-2009 has resulted in narrower confidence intervals for estimates of F, SSB, and catch components (landings, discards and total removals). Some uncertainty remains over the unallocated component of removals and how this could be divided between removals caused by natural mortality and removals related to fishing (for example, escape mortality and area misreporting).

## Weights-at-age

In this assessment, simple linear growth models have been fitted to cohort weights-at-age data and used to generate weights-at-age in the forecast. These models fit reasonably well, but this approach is quite simplistic and may be missing important growth characteristics such as variable growth within a cohort. This may lead to greater uncertainty in the forecast

## Model formulation

Models such as the modified TSA used this year, based largely on survey data, are becoming the de facto standard in several ICES assessments for which problems have existed with commercial catch data (see this report, and also WGNSSK 2006). Other examples include BADAPT and SURBA. While these types of models are essential in order to address data problems, it needs to be borne in mind that there are two main problems with such approaches. Firstly, survey data are based on far fewer samples, and are therefore more variable than catch data. It is therefore likely that precision is sacrificed (to a certain extent) to reduce bias. Secondly, a survey-based assessment estimates removals from the stock and total mortality, rather than landings and fishing mortality, and is therefore more difficult to use as the basis of quota advice than corresponding catch-based approaches. It is therefore thought to be appropriate to reinclude catch data when they become more reliable, and investigations have indicated that this has been the case in the years 2006-2009.

## Stock connectivity

There is uncertainty concerning the stock definition and hence the degree of connectivity between the VIa haddock stock and the North Sea haddock stock. Since these stocks are currently assessed separately, it is possible that the two stock assessments are both affected by uncertainties in catch data relating to area misreporting.

### 3.3.9 Recommendations for next benchmark

Some ways of addressing these issues are proposed here. All aspects are considered important and the proposed time frame would be to work on these in 2011-2012. Continuing the work on management plan development is also important.

## Landings and discards

There should be a full analysis of the precision and bias of catch-at-age data. Although catch data between 2006-2009 are thought to represent a large proportion of the true catch, further analysis would help to put a clearer estimate on the uncertainty of this. Measures such as the UK Registration of Buyers and Sellers legislation seem to have greatly improved the reliability of commercial landings data for the last three years. The reported catches from 2006 onwards are within the bounds of error of the estimated catch.

## Effort

A VMS-based analysis of lpue could help to address the concern that currently commercial cpue or lpue data cannot be used in the assessment. With the increased requirement for vessels to operate with VMS it is likely that the quality of effort data will improve. This will lead to an improved time-series of effort data in the future but still leaves the uncertainties regarding the earlier years in the time-series.

## Surveys

As the time-series lengthens, an analysis of the new IGFS should take place in order to check the quality and consistency of this survey for tuning purposes and hence decide on its inclusion in the assessment.

## Weights-at-age

The growth characteristics of this haddock stock are very variable, and seem to be strongly driven by cohort effects rather than year effects: that is, early life-history events determine the subsequent growth potential of each cohort. Work is underway at Marine Scotland (Aberdeen) and elsewhere to develop improved models of growth, and it is hoped that these will improve stock forecasts in the future. Consideration of using stock weights from the survey, instead of the estimated weights-atage could also be addressed at a benchmark assessment.

## Other modelling

Growth modelling could help with forecasts of mean weights-at-age. It may also be of interest to use bioeconomic models to address questions to do with feedbacks between quota, uptake of quota and strong drivers of quota uptake and fishers' behaviour, for example, fuel price.

Other assessment models could be considered where information from the age structure of the catch data could be incorporated in the assessment for the years where the catch data are currently excluded (1995-2005).

The WG recommends that this stock should be benchmarked with the North Sea haddock stock in 2011.

### 3.3.10 Management considerations

This stock is at a low level of biomass, with recruitment impaired in four of the five
most recent years. An agreed long-term management plan, which takes into account the recruitment characteristics of this stock, is needed.

The fishery for haddock is limited to vessels from mainly Scotland and Ireland. Uptake of quota was low, and hence TAC was not limiting, in both 2008 and 2009. Discarding, however, remained high in these two years, accounting for $32 \%$ of estimated total catch in 2008 and $39 \%$ of estimated total catch in 2009.

Reallocation of effort from Division VIa into other ICES areas and switching between mesh categories may also be significant. While there has been a general decline in the haddock fishery in Division VIa, both Irish and Scottish sources suggest that there is an increasing focus in the corresponding Division VIb (Rockall) fishery. In addition, a few Scottish fishermen are testing the viability of using paired gear (both seine and trawl) at Rockall: if this proves successful, then there is the distinct possibility that effective effort in Division VIb will increase considerably. This fishery is particularly attractive given the lack of effort restrictions in the area.

Table 3.3.1. Haddock in Division VIa. Nominal landings, as officially reported to ICES and estimated by the WG.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 8 | 9 | - | 9 | 1 | 7 | 1 | - | 1 | 3 | 2 | 2 | 1 |
| Denmark | + | + | + | + | - | 1 | - | 1 | 1 | - | - | - | - |
| Faroe Is | - | 13 | - | 1 | - | - | - | - | - | - | - | - | - |
| France | 3001 | 13352 | 8632 | 7612 | 761 | 1132 | 753 | 671 | 445 | 270 | 394 | 788 | 282 |
| Germany | 4 | 4 | 15 | 1 | 2 | 9 | 19 | 14 | 2 | 1 | 1 | 2 | 1 |
| Ireland | 2731 | 2171 | 773 | 710 | 700 | 911 | 746 | 1406 | 1399 | 1447 | 1352 | 1054 | 677 |
| Norway | 54 | 74 | 46 | 12 | 72 | 40 | 7 | 13 | 16 | 21 | 28 | 18 | 70 |
| Spain | - | - | - | - | - | - | - | 1 | - | - | 2 | 4 | - |
| UK (E\&W) ${ }^{3}$ | 114 | 235 | 164 | 137 | 132 | 155 | 254 | 322 | 448 | 493 | 458 | 315 | 199 |
| UK (N. Ire) | 35 |  |  |  |  |  |  |  | $\ldots$ | ... | $\cdots$ | ... | ... |
| UK (Scot.) | 15151 | 19940 | 10964 | 8434 | 5263 | 10423 | 7421 | 10367 | 10790 | 10352 | 12125 | 8630 | 5933 |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Total reported | 21098 | 23781 | 12825 | 10065 | 6932 | 12678 | 9201 | 12794 | 13102 | 12587 | 14360 | 10813 | 7163 |
| WG estimates | 21136 | 16688 | 10135 | 10557 | 11350 | 19060 | 14243 | 12368 | 13453 | 12874 | 14401 | 10430 | 6952 |

## 1Preliminary.

2Includes Divisions $\mathrm{Vb}(E C)$ and VIb.
31989-2005 N. Ireland included with England and Wales.
$\mathrm{n} / \mathrm{a}=$ Not available.
WG estimates refers to the sum-of-products of landings and weights-at-age provided to the WG, rather than the estimated removals produced in the final assessment.

Table 3.3.1. Continued. Haddock in Division VIa. Nominal landings, as officially reported to ICES and estimated by the WG.

| Country | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 20091 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 2 | - | - | + | - | - | - | - | - |
| Denmark | - | - | + | - |  |  |  | - | - |
| Faroe Is. | - | - | - | 4 | - | 1 | 2 | - | 1 |
| France | 160 | 151 | 183 | 173 | 273 | 291 | 211 | 85 | 124 |
| Germany | 1 | + | - | - | 1 | 7 | - | 1 | - |
| Ireland | 744 | 672 | 497 | 194 | 152 | 526 | 759 | 879 | 297 |
| Norway | 32 | 30 | 23 | 4 | 21 | 17 | 16 | 28 | 18 |
| Spain | 4 | 4 | 5 | - | 47 | 44 | 5 | - | - |
| UK (E\&W)3 | 201 | 237 | 107 | 93 | 42 | 19 | 193 | - | 2 |
| UK (N. Ire) | ... | ... | ... | ... | ... | ... | ... | - | 8 |
| UK (Scot.) | 5886 | 5988 | 4582 | 2909 | 2025 | 4928 | 2587 | - | 2351 |
| UK (total) |  |  |  |  |  |  | - | 1769 | 2380 |
| Netherlands | - | - | - | 1 | - | - | - | - | - |
| Total reported | 7030 | 7082 | 5397 | 3378 | 2561 | 5833 | 3773 | 2762 | 2695 |
| WG estimates | 6731 | 7097 | 5334 | 3199 | 3148 | 5723 | 3702 | 2801 | 2800 |

## 1Preliminary.

2Includes Divisions $\operatorname{Vb}(E C)$ and VIb.
31989-2005 N. Ireland included with England and Wales.
$\mathbf{n} / \mathbf{a}=$ Not available.
WG estimates refers to the sum-of-products of landings and weights-at-age provided to the WG, rather than the estimated removals produced in the final assessment.

Table 3.3.2. Haddock in Division VIa. Total catch-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 451 | 1059 | 1341 | 72461 | 6816 | 294 | 274 | 174 | 11 |
| 1966 | 5953 | 1595 | 529 | 1113 | 47431 | 1926 | 64 | 32 | 57 |
| 1967 | 40122 | 19185 | 19332 | 951 | 265 | 24979 | 400 | 9 | 14 |
| 1968 | 27 | 129418 | 38393 | 3079 | 356 | 681 | 14063 | 727 | 43 |
| 1969 | 2742 | 84 | 160706 | 10260 | 1434 | 268 | 379 | 4576 | 191 |
| 1970 | 17189 | 6317 | 519 | 95114 | 2770 | 173 | 89 | 145 | 585 |
| 1971 | 6604 | 71481 | 3915 | 3328 | 79966 | 545 | 127 | 7 | 20 |
| 1972 | 14215 | 20713 | 85141 | 2718 | 2336 | 53823 | 504 | 50 | 19 |
| 1973 | 19589 | 47387 | 16907 | 19477 | 258 | 1222 | 33193 | 150 | 32 |
| 1974 | 63698 | 68837 | 11562 | 10757 | 6317 | 83 | 447 | 11463 | 104 |
| 1975 | 6849 | 179349 | 34957 | 3339 | 3350 | 1882 | 95 | 98 | 3454 |
| 1976 | 4227 | 24337 | 72330 | 15224 | 1588 | 1491 | 868 | 21 | 7 |
| 1977 | 4552 | 13109 | 3468 | 35948 | 5705 | 680 | 495 | 308 | 28 |
| 1978 | 57 | 15942 | 2095 | 971 | 24357 | 2938 | 351 | 247 | 338 |
| 1979 | 5697 | 70070 | 17282 | 1865 | 470 | 9863 | 833 | 114 | 145 |
| 1980 | 13 | 22729 | 21927 | 5636 | 922 | 143 | 3082 | 229 | 22 |
| 1981 | 764 | 251 | 83911 | 20697 | 1768 | 194 | 39 | 822 | 39 |
| 1982 | 136 | 15492 | 5019 | 73676 | 8167 | 898 | 108 | 272 | 288 |
| 1983 | 2084 | 14524 | 20233 | 6040 | 36122 | 3398 | 597 | 41 | 194 |
| 1984 | 269 | 98976 | 8626 | 12910 | 6242 | 22790 | 2449 | 371 | 43 |
| 1985 | 155 | 22820 | 78922 | 4667 | 4184 | 1789 | 11189 | 964 | 84 |
| 1986 | 2979 | 8127 | 11235 | 45367 | 1823 | 916 | 449 | 2611 | 344 |
| 1987 | 1498 | 89021 | 16824 | 10150 | 23857 | 1452 | 1116 | 642 | 1818 |
| 1988 | 7582 | 10007 | 58414 | 7598 | 4185 | 9255 | 428 | 235 | 177 |
| 1989 | 3773 | 5010 | 3420 | 25724 | 2755 | 1556 | 3634 | 255 | 84 |
| 1990 | 437 | 37247 | 5856 | 1884 | 12158 | 871 | 279 | 519 | 48 |
| 1991 | 8921 | 36924 | 21991 | 1259 | 834 | 5132 | 412 | 283 | 410 |
| 1992 | 4332 | 51840 | 18971 | 11331 | 565 | 236 | 1577 | 157 | 37 |
| 1993 | 2196 | 43659 | 60785 | 20763 | 4669 | 306 | 219 | 915 | 70 |
| 1994 | 2843 | 19484 | 32638 | 21527 | 5671 | 1579 | 76 | 175 | 237 |
| 1995 | 7692 | 17580 | 15759 | 23599 | 6865 | 1472 | 387 | 34 | 111 |
| 1996 | 10249 | 33344 | 39812 | 6641 | 10225 | 3663 | 1007 | 324 | 23 |
| 1997 | 2984 | 23843 | 10507 | 21550 | 2178 | 2668 | 870 | 259 | 59 |
| 1998 | 2058 | 11421 | 18001 | 8032 | 15116 | 1352 | 1036 | 377 | 124 |
| 1999 | 6898 | 6179 | 18055 | 11569 | 3004 | 4919 | 579 | 452 | 96 |
| 2000 | 5709 | 50142 | 6642 | 8596 | 4213 | 1055 | 1104 | 205 | 133 |
| 2001 | 11818 | 11023 | 33496 | 2432 | 3666 | 1521 | 533 | 314 | 65 |
| 2002 | 1362 | 16427 | 12394 | 32248 | 833 | 714 | 549 | 238 | 144 |
| 2003 | 3861 | 6972 | 5592 | 6848 | 12830 | 222 | 209 | 70 | 34 |
| 2004 | 2727 | 15159 | 6506 | 2384 | 3839 | 6706 | 286 | 101 | 26 |
| 2005 | 3965 | 7190 | 6202 | 3700 | 2116 | 2669 | 2704 | 57 | 42 |
| 2006 | 817 | 16031 | 4831 | 3844 | 3801 | 3109 | 2731 | 2750 | 33 |
| 2007 | 257 | 1777 | 15850 | 2897 | 1725 | 2428 | 811 | 904 | 478 |
| 2008 | 1840 | 2409 | 2330 | 4421 | 587 | 609 | 868 | 255 | 185 |
| 2009 | 2012 | 4977 | 433 | 427 | 6651 | 510 | 334 | 253 | 78 |

Table 3.3.2. Continued. Haddock in Division VIa. Total catch-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 6 | 6 | 0 | 0 | 0 | 0 | 0 | 24 |
| 1966 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 |
| $1967$ | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 19 |
| 1968 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 52 |
| 1969 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 200 |
| 1970 | 13 | 2 | 0 | 0 | 0 | 0 | 0 | 600 |
| 1971 | 175 | 16 | 0 | 0 | 0 | 0 | 0 | 212 |
| 1972 | 0 | 67 | 0 | 0 | 0 | 0 | 0 | 86 |
| 1973 | 6 | 125 | 0 | 0 | 0 | 0 | 0 | 163 |
| 1974 | 34 | 31 | 0 | 1 | 4 | 0 | 0 | 174 |
| 1975 | 72 | 8 | 0 | 0 | 0 | 0 | 0 | 3534 |
| 1976 | 1103 | 4 | 0 | 5 | 0 | 0 | 0 | 1119 |
| 1977 | 11 | 259 | 5 | 0 | 0 | 0 | 0 | 304 |
| 1978 | 7 | 17 | 211 | 3 | 0 | 0 | 0 | 575 |
| 1979 | 28 | 3 | 1 | 42 | 1 | 0 | 0 | 221 |
| $1980$ | 5 | 21 | 3 | 0 | 4 | 0 | 0 | 54 |
| $1981$ | 14 | 2 | 2 | 1 | 0 | 1 | 0 | 60 |
| 1982 | 31 | 12 | 1 | 0 | 0 | 0 | 0 | 332 |
| 1983 | 195 | 40 | 15 | 0 | 0 | 0 | 0 | 444 |
| $1984$ | 44 | 73 | 3 | 0 | 0 | 0 | 0 | 162 |
| 1985 | 4 | 8 | 56 | 4 | 0 | 0 | 1 | 157 |
| 1986 | 38 | 7 | 15 | 1 | 3 | 0 | 0 | 409 |
| $1987$ | 326 | 20 | 15 | 9 | 3 | 12 | 0 | 2203 |
| $1988$ | 935 | 45 | 3 | 1 | 3 | 2 | 0 | 1167 |
| 1989 | 87 | 437 | 56 | 1 | 1 | 0 | 0 | 666 |
| 1990 | 22 | 12 | 2 | 0 | 0 | 0 | 0 | 85 |
| $1991$ | 24 | 11 | 5 | 6 | 0 | 0 | 1 | 457 |
| 1992 | 108 | 25 | 0 | 0 | 0 | 0 | 0 | 169 |
| 1993 | 107 | 44 | 25 | 1 | 2 | 0 | 0 | 250 |
| $1994$ | 17 | 16 | 9 | 1 | 0 | 0 | 0 | 279 |
| 1995 | 90 | 2 | 0 | 0 | 0 | 0 | 0 | 203 |
| 1996 | 40 | 12 | 4 | 0 | 0 | 0 | 0 | 80 |
| 1997 | 1 | 7 | 1 | 0 | 0 | 0 | 0 | 67 |
| $1998$ | 45 | 2 | 4 | 1 | 0 | 0 | 0 | 175 |
| 1999 | 12 | 2 | 1 | 2 | 1 | 0 | 0 | 115 |
| 2000 | 21 | 1 | 0 | 0 | 0 | 0 | 0 | 156 |
| 2001 | 25 | 11 | 0 | 3 | 0 | 0 | 0 | 104 |
| 2002 | 18 | 9 | 0 | 0 | 0 | 0 | 0 | 172 |
| 2003 | 12 | 10 | 0 | 0 | 0 | 0 | 0 | 56 |
| 2004 | 6 | 2 | 2 | 0 | 0 | 0 | 0 | 37 |
| 2005 | 5 | 1 | 1 | 0 | 0 | 0 | 0 | 48 |
| 2006 | 26 | 5 | 0 | 0 | 1 | 0 | 0 | 65 |
| 2007 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 485 |
| 2008 | 122 | 0 | 0 | 0 | 0 | 0 | 0 | 307 |
| 2009 | 41 | 31 | 0 | 0 | 0 | 0 | 0 | 151 |

Table 3.3.3. Haddock in Division VIa. Landings-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 0 | 33 | 463 | 60967 | 6753 | 294 | 274 | 174 | 11 |
| 1966 | 0 | 58 | 175 | 1082 | 46902 | 1926 | 64 | 32 | 57 |
| 1967 | 0 | 595 | 6136 | 782 | 262 | 24979 | 400 | 9 | 14 |
| 1968 | 0 | 3665 | 12439 | 2573 | 354 | 681 | 14063 | 727 | 43 |
| 1969 | 0 | 3 | 45819 | 8766 | 1423 | 268 | 379 | 4576 | 191 |
| 1970 | 0 | 169 | 170 | 78402 | 2747 | 173 | 89 | 145 | 585 |
| 1971 | 0 | 1925 | 1149 | 2665 | 78909 | 545 | 127 | 7 | 20 |
| 1972 | 0 | 576 | 26700 | 2225 | 2312 | 53823 | 504 | 50 | 19 |
| 1973 | 0 | 1252 | 5301 | 16109 | 256 | 1222 | 33193 | 150 | 32 |
| 1974 | 0 | 1706 | 3318 | 8625 | 6261 | 83 | 447 | 11463 | 104 |
| 1975 | 0 | 4629 | 10534 | 2735 | 3315 | 1882 | 95 | 98 | 3454 |
| 1976 | 0 | 745 | 22563 | 12358 | 1571 | 1491 | 868 | 21 | 7 |
| 1977 | 0 | 451 | 1317 | 29456 | 5645 | 680 | 495 | 308 | 28 |
| 1978 | 0 | 1030 | 1006 | 813 | 23620 | 2912 | 344 | 247 | 338 |
| 1979 | 0 | 2068 | 10448 | 1761 | 468 | 9810 | 833 | 114 | 145 |
| 1980 | 0 | 2505 | 12871 | 5341 | 915 | 143 | 3082 | 229 | 22 |
| 1981 | 0 | 200 | 20553 | 15695 | 1768 | 194 | 39 | 822 | 39 |
| 1982 | 0 | 250 | 1342 | 46283 | 8004 | 898 | 108 | 272 | 288 |
| 1983 | 0 | 568 | 4917 | 4585 | 34659 | 3387 | 597 | 41 | 194 |
| 1984 | 0 | 3341 | 4386 | 10754 | 5959 | 20352 | 2449 | 371 | 43 |
| 1985 | 0 | 939 | 19434 | 4437 | 4112 | 1782 | 11031 | 964 | 84 |
| 1986 | 0 | 603 | 4812 | 26770 | 1823 | 916 | 449 | 2611 | 344 |
| 1987 | 0 | 4254 | 7388 | 9206 | 23551 | 1452 | 1116 | 642 | 1818 |
| 1988 | 0 | 847 | 20687 | 6873 | 4091 | 9205 | 428 | 235 | 177 |
| 1989 | 0 | 927 | 1414 | 18417 | 2744 | 1556 | 3633 | 255 | 84 |
| 1990 | 0 | 787 | 3198 | 1342 | 9450 | 848 | 279 | 519 | 48 |
| 1991 | 0 | 2145 | 10578 | 1217 | 834 | 5131 | 412 | 283 | 410 |
| 1992 | 0 | 691 | 10194 | 10010 | 553 | 236 | 1575 | 157 | 37 |
| 1993 | 0 | 745 | 15008 | 15975 | 4594 | 290 | 219 | 910 | 70 |
| 1994 | 0 | 1017 | 6326 | 15037 | 5240 | 1484 | 76 | 175 | 237 |
| 1995 | 0 | 540 | 3669 | 12774 | 6483 | 1472 | 387 | 34 | 111 |
| 1996 | 0 | 437 | 9457 | 4968 | 8626 | 3622 | 1007 | 324 | 23 |
| 1997 | 0 | 883 | 2831 | 16921 | 2125 | 2638 | 870 | 259 | 59 |
| 1998 | 0 | 1345 | 7129 | 5675 | 13387 | 1352 | 1036 | 377 | 124 |
| 1999 | 0 | 346 | 5501 | 7159 | 2960 | 4864 | 493 | 452 | 96 |
| 2000 | 0 | 759 | 2507 | 5864 | 3841 | 1054 | 1090 | 205 | 133 |
| 2001 | 0 | 245 | 8535 | 1822 | 3523 | 1393 | 533 | 314 | 65 |
| 2002 | 0 | 177 | 1227 | 13557 | 691 | 707 | 549 | 199 | 144 |
| 2003 | 0 | 21 | 1029 | 2150 | 8809 | 221 | 206 | 69 | 34 |
| 2004 | 0 | 14 | 245 | 804 | 1819 | 4071 | 286 | 100 | 26 |
| 2005 | 0 | 7 | 287 | 792 | 1252 | 1212 | 2018 | 57 | 42 |
| 2006 | 0 | 67 | 567 | 1513 | 2300 | 2504 | 2259 | 2192 | 33 |
| 2007 | 0 | 34 | 842 | 1121 | 1429 | 2394 | 778 | 855 | 478 |
| 2008 | 0 | 21 | 297 | 2718 | 546 | 584 | 752 | 254 | 161 |
| 2009 | 0 | 4 | 57 | 188 | 3912 | 485 | 286 | 207 | 78 |

Table 3.3.3. Continued. Haddock in Division VIa. Landings-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 6 | 6 | 0 | 0 | 0 | 0 | 0 | 24 |
| 1966 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 |
| 1967 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 19 |
| 1968 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 52 |
| 1969 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 200 |
| 1970 | 13 | 2 | 0 | 0 | 0 | 0 | 0 | 600 |
| 1971 | 175 | 16 | 0 | 0 | 0 | 0 | 0 | 212 |
| 1972 | 0 | 67 | 0 | 0 | 0 | 0 | 0 | 86 |
| 1973 | 6 | 125 | 0 | 0 | 0 | 0 | 0 | 163 |
| 1974 | 34 | 31 | 0 | 1 | 4 | 0 | 0 | 174 |
| 1975 | 72 | 8 | 0 | 0 | 0 | 0 | 0 | 3534 |
| 1976 | 1103 | 4 | 0 | 5 | 0 | 0 | 0 | 1119 |
| 1977 | 11 | 259 | 5 | 0 | 0 | 0 | 0 | 304 |
| 1978 | 7 | 17 | 211 | 3 | 0 | 0 | 0 | 575 |
| 1979 | 28 | 3 | 1 | 42 | 1 | 0 | 0 | 221 |
| 1980 | 5 | 21 | 3 | 0 | 4 | 0 | 0 | 54 |
| 1981 | 14 | 2 | 2 | 1 | 0 | 1 | 0 | 60 |
| 1982 | 31 | 12 | 1 | 0 | 0 | 0 | 0 | 332 |
| 1983 | 195 | 40 | 15 | 0 | 0 | 0 | 0 | 444 |
| 1984 | 44 | 73 | 3 | 0 | 0 | 0 | 0 | 162 |
| 1985 | 4 | 8 | 56 | 4 | 0 | 0 | 1 | 157 |
| 1986 | 38 | 7 | 15 | 1 | 3 | 0 | 0 | 409 |
| $1987$ | 326 | 20 | 15 | 9 | 3 | 12 | 0 | 2203 |
| 1988 | 935 | 45 | 3 | 1 | 3 | 2 | 0 | 1167 |
| 1989 | 87 | 437 | 56 | 1 | 1 | 0 | 0 | 666 |
| 1990 | 22 | 12 | 2 | 0 | 0 | 0 | 0 | 85 |
| 1991 | 24 | 11 | 5 | 6 | 0 | 0 | 1 | 457 |
| 1992 | 108 | 25 | 0 | 0 | 0 | 0 | 0 | 169 |
| 1993 | 107 | 44 | 25 | 1 | 2 | 0 | 0 | 250 |
| 1994 | 17 | 16 | 9 | 1 | 0 | 0 | 0 | 279 |
| 1995 | 90 | 2 | 0 | 0 | 0 | 0 | 0 | 203 |
| 1996 | 40 | 12 | 4 | 0 | 0 | 0 | 0 | 80 |
| 1997 | 1 | 7 | 1 | 0 | 0 | 0 | 0 | 67 |
| $1998$ | 45 | 2 | 4 | 1 | 0 | 0 | 0 | 175 |
| 1999 | 12 | 2 | 1 | 2 | 1 | 0 | 0 | 115 |
| 2000 | 21 | 1 | 0 | 0 | 0 | 0 | 0 | 156 |
| 2001 | 25 | 11 | 0 | 3 | 0 | 0 | 0 | 104 |
| 2002 | 18 | 9 | 0 | 0 | 0 | 0 | 0 | 172 |
| 2003 | 11 | 10 | 0 | 0 | 0 | 0 | 0 | 55 |
| 2004 | 6 | 2 | 2 | 0 | 0 | 0 | 0 | 37 |
| 2005 | 5 | 1 | 1 | 0 | 0 | 0 | 0 | 48 |
| 2006 | 26 | 5 | 0 | 0 | 1 | 0 | 0 | 65 |
| 2007 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 485 |
| 2008 | 122 | 0 | 0 | 0 | 0 | 0 | 0 | 283 |
| 2009 | 41 | 31 | 0 | 0 | 0 | 0 | 0 | 151 |

Table 3.3.4. Haddock in Division VIa. Discards-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 451 | 1026 | 877 | 11494 | 63 | 0 | 0 | 0 | 0 |
| 1966 | 5953 | 1537 | 354 | 31 | 529 | 0 | 0 | 0 | 0 |
| 1967 | 40122 | 18590 | 13196 | 169 | 3 | 0 | 0 | 0 | 0 |
| 1968 | 27 | 125753 | 25954 | 506 | 3 | 0 | 0 | 0 | 0 |
| 1969 | 2742 | 81 | 114887 | 1493 | 11 | 0 | 0 | 0 | 0 |
| 1970 | 17189 | 6148 | 348 | 16712 | 23 | 0 | 0 | 0 | 0 |
| 1971 | 6604 | 69556 | 2766 | 663 | 1057 | 0 | 0 | 0 | 0 |
| 1972 | 14215 | 20137 | 58442 | 494 | 24 | 0 | 0 | 0 | 0 |
| 1973 | 19589 | 46135 | 11607 | 3368 | 2 | 0 | 0 | 0 | 0 |
| 1974 | 63698 | 67131 | 8244 | 2132 | 56 | 0 | 0 | 0 | 0 |
| 1975 | 6849 | 174721 | 24423 | 604 | 35 | 0 | 0 | 0 | 0 |
| 1976 | 4227 | 23593 | 49767 | 2866 | 17 | 0 | 0 | 0 | 0 |
| 1977 | 4552 | 12658 | 2152 | 6492 | 59 | 0 | 0 | 0 | 0 |
| 1978 | 55 | 14911 | 1090 | 157 | 738 | 27 | 7 | 0 | 0 |
| 1979 | 5697 | 68002 | 6833 | 104 | 2 | 53 | 0 | 0 | 0 |
| 1980 | 13 | 20224 | 9057 | 295 | 7 | 0 | 0 | 0 | 0 |
| 1981 | 764 | 51 | 63359 | 5002 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 136 | 15241 | 3678 | 27393 | 163 | 0 | 0 | 0 | 0 |
| 1983 | 2084 | 13957 | 15316 | 1456 | 1464 | 12 | 0 | 0 | 0 |
| 1984 | 269 | 95634 | 4240 | 2156 | 284 | 2438 | 0 | 0 | 0 |
| 1985 | 155 | 21882 | 59488 | 231 | 71 | 6 | 159 | 0 | 0 |
| 1986 | 2979 | 7524 | 6423 | 18597 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1498 | 84767 | 9436 | 944 | 306 | 0 | 0 | 0 | 0 |
| 1988 | 7582 | 9160 | 37727 | 725 | 95 | 49 | 0 | 0 | 0 |
| 1989 | 3773 | 4083 | 2007 | 7308 | 11 | 0 | 1 | 0 | 0 |
| 1990 | 437 | 36460 | 2658 | 542 | 2708 | 23 | 0 | 0 | 0 |
| 1991 | 8921 | 34779 | 11413 | 42 | 0 | 1 | 0 | 0 | 0 |
| 1992 | 4331 | 51148 | 8776 | 1322 | 12 | 0 | 2 | 0 | 0 |
| 1993 | 2196 | 42914 | 45777 | 4787 | 74 | 16 | 0 | 5 | 0 |
| 1994 | 2843 | 18467 | 26312 | 6490 | 432 | 94 | 0 | 0 | 0 |
| 1995 | 7692 | 17040 | 12090 | 10825 | 382 | 0 | 0 | 0 | 0 |
| 1996 | 10249 | 32907 | 30354 | 1674 | 1599 | 41 | 0 | 0 | 0 |
| 1997 | 2984 | 22961 | 7676 | 4629 | 53 | 30 | 0 | 0 | 0 |
| 1998 | 2058 | 10075 | 10872 | 2357 | 1728 | 0 | 0 | 0 | 0 |
| 1999 | 6898 | 5834 | 12554 | 4410 | 44 | 54 | 86 | 0 | 0 |
| 2000 | 5709 | 49383 | 4136 | 2731 | 372 | 1 | 14 | 0 | 0 |
| 2001 | 11818 | 10778 | 24961 | 611 | 143 | 128 | 0 | 0 | 0 |
| 2002 | 1362 | 16250 | 11168 | 18692 | 142 | 8 | 0 | 39 | 0 |
| 2003 | 3861 | 6951 | 4564 | 4697 | 4021 | 2 | 2 | 1 | 0 |
| 2004 | 2727 | 15146 | 6261 | 1580 | 2021 | 2635 | 0 | 1 | 0 |
| 2005 | 3965 | 7184 | 5915 | 2908 | 864 | 1457 | 686 | 0 | 1 |
| 2006 | 817 | 15964 | 4263 | 2331 | 1501 | 605 | 471 | 557 | 0 |
| 2007 | 257 | 1743 | 15008 | 1775 | 296 | 34 | 33 | 48 | 0 |
| 2008 | 1840 | 2388 | 2033 | 1703 | 41 | 25 | 116 | 1 | 24 |
| 2009 | 2012 | 4972 | 376 | 239 | 2740 | 25 | 48 | 46 | 0 |

Table 3.3.4. Continued. Haddock in Division VIa. Discards-at-age numbers (000s). Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1966 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1967 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1968 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1969 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 24 |
| 2009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 3.3.5. Haddock in Division VIa. Weights-at-age (kg) in total catch. Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 0.04 | 0.16 | 0.242 | 0.412 | 0.692 | 0.916 | 1.041 | 1.249 | 1.517 |
| 1966 | 0.04 | 0.162 | 0.251 | 0.555 | 0.572 | 1.041 | 1.125 | 1.325 | 1.522 |
| 1967 | 0.04 | 0.16 | 0.266 | 0.569 | 0.573 | 0.667 | 1.177 | 1.844 | 1.611 |
| 1968 | 0.04 | 0.159 | 0.264 | 0.567 | 0.823 | 0.731 | 0.811 | 1.43 | 1.903 |
| 1969 | 0.04 | 0.158 | 0.243 | 0.526 | 0.916 | 1.042 | 1.024 | 0.999 | 1.569 |
| 1970 | 0.04 | 0.161 | 0.23 | 0.368 | 0.812 | 1.283 | 1.262 | 1.043 | 1.342 |
| 1971 | 0.04 | 0.16 | 0.248 | 0.341 | 0.546 | 1.04 | 1.313 | 1.651 | 1.426 |
| 1972 | 0.04 | 0.16 | 0.249 | 0.38 | 0.53 | 0.546 | 0.984 | 1.499 | 1.538 |
| 1973 | 0.04 | 0.159 | 0.251 | 0.384 | 0.597 | 0.512 | 0.571 | 1.185 | 1.706 |
| 1974 | 0.04 | 0.159 | 0.248 | 0.368 | 0.527 | 0.764 | 0.685 | 0.798 | 1.142 |
| 1975 | 0.04 | 0.159 | 0.26 | 0.428 | 0.581 | 0.832 | 1.027 | 1.001 | 1.009 |
| 1976 | 0.04 | 0.159 | 0.256 | 0.459 | 0.592 | 0.831 | 1.095 | 1.585 | 1.084 |
| 1977 | 0.04 | 0.161 | 0.274 | 0.406 | 0.684 | 0.8 | 1.128 | 1.337 | 1.117 |
| 1978 | 0.068 | 0.134 | 0.278 | 0.388 | 0.516 | 0.827 | 1.045 | 1.152 | 1.399 |
| 1979 | 0.032 | 0.182 | 0.325 | 0.457 | 0.73 | 0.777 | 1.04 | 1.491 | 1.944 |
| 1980 | 0.077 | 0.134 | 0.319 | 0.572 | 0.719 | 0.998 | 0.985 | 1.143 | 1.565 |
| 1981 | 0.082 | 0.252 | 0.245 | 0.467 | 0.887 | 0.975 | 1.376 | 1.294 | 1.347 |
| 1982 | 0.038 | 0.157 | 0.273 | 0.376 | 0.746 | 1.126 | 1.539 | 1.549 | 1.514 |
| 1983 | 0.05 | 0.178 | 0.282 | 0.461 | 0.557 | 1.002 | 1.37 | 1.716 | 1.558 |
| 1984 | 0.059 | 0.149 | 0.319 | 0.456 | 0.688 | 0.667 | 1.087 | 1.392 | 2.075 |
| 1985 | 0.019 | 0.138 | 0.268 | 0.486 | 0.636 | 0.802 | 0.868 | 1.272 | 1.277 |
| 1986 | 0.064 | 0.182 | 0.27 | 0.362 | 0.637 | 0.903 | 1.115 | 1.043 | 1.418 |
| 1987 | 0.028 | 0.168 | 0.27 | 0.418 | 0.566 | 0.88 | 1.105 | 1.25 | 1.147 |
| 1988 | 0.085 | 0.17 | 0.254 | 0.444 | 0.562 | 0.704 | 1.027 | 1.28 | 1.279 |
| 1989 | 0.052 | 0.226 | 0.301 | 0.402 | 0.625 | 0.749 | 0.894 | 1.115 | 1.465 |
| 1990 | 0.073 | 0.112 | 0.355 | 0.445 | 0.534 | 0.891 | 1.108 | 1.28 | 1.823 |
| 1991 | 0.058 | 0.184 | 0.297 | 0.547 | 0.618 | 0.678 | 0.931 | 1.053 | 1.091 |
| 1992 | 0.05 | 0.133 | 0.321 | 0.437 | 0.766 | 0.892 | 0.932 | 1.407 | 1.493 |
| 1993 | 0.037 | 0.108 | 0.277 | 0.458 | 0.65 | 0.861 | 0.898 | 1.022 | 1.514 |
| 1994 | 0.031 | 0.169 | 0.253 | 0.405 | 0.611 | 0.698 | 0.929 | 0.959 | 0.909 |
| 1995 | 0.03 | 0.149 | 0.274 | 0.354 | 0.553 | 0.833 | 0.978 | 1.322 | 1.059 |
| 1996 | 0.047 | 0.128 | 0.243 | 0.404 | 0.462 | 0.645 | 0.75 | 0.754 | 1.122 |
| 1997 | 0.048 | 0.153 | 0.263 | 0.394 | 0.614 | 0.73 | 0.925 | 1.057 | 0.921 |
| 1998 | 0.089 | 0.164 | 0.283 | 0.382 | 0.502 | 0.689 | 0.802 | 0.951 | 1.006 |
| 1999 | 0.035 | 0.172 | 0.255 | 0.365 | 0.494 | 0.611 | 0.729 | 0.84 | 1.067 |
| 2000 | 0.053 | 0.127 | 0.27 | 0.361 | 0.447 | 0.572 | 0.719 | 0.84 | 0.749 |
| 2001 | 0.05 | 0.112 | 0.242 | 0.403 | 0.432 | 0.514 | 0.657 | 0.808 | 1.029 |
| 2002 | 0.048 | 0.118 | 0.208 | 0.307 | 0.521 | 0.606 | 0.632 | 0.636 | 0.81 |
| 2003 | 0.036 | 0.124 | 0.239 | 0.282 | 0.382 | 0.652 | 0.648 | 0.908 | 0.945 |
| 2004 | 0.033 | 0.112 | 0.189 | 0.29 | 0.313 | 0.373 | 0.541 | 0.715 | 0.782 |
| 2005 | 0.053 | 0.103 | 0.198 | 0.295 | 0.451 | 0.429 | 0.525 | 1.163 | 0.916 |
| 2006 | 0.024 | 0.155 | 0.254 | 0.326 | 0.388 | 0.471 | 0.496 | 0.563 | 1.242 |
| 2007 | 0.060 | 0.115 | 0.219 | 0.331 | 0.404 | 0.456 | 0.550 | 0.593 | 0.682 |
| 2008 | 0.022 | 0.113 | 0.245 | 0.367 | 0.492 | 0.570 | 0.619 | 0.708 | 0.770 |
| 2009 | 0.048 | 0.135 | 0.252 | 0.357 | 0.410 | 0.570 | 0.633 | 0.630 | 0.897 |

Table 3.3.5. Continued. Haddock in Division VIa. Weights-at-age (kg) in total catch. Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 1.92 | 1.833 | 0 | 0 | 0 | 0 | 0 | 1.713 |
| 1966 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.522 |
| 1967 | 2.355 | 0 | 0 | 0 | 0 | 0 | 0 | 1.786 |
| 1968 | 2.516 | 0 | 0 | 0 | 0 | 0 | 0 | 2.005 |
| 1969 | 2.065 | 0 | 0 | 0 | 0 | 0 | 0 | 1.590 |
| 1970 | 1.791 | 1.213 | 0 | 0 | 0 | 0 | 0 | 1.352 |
| 1971 | 1.466 | 2.042 | 0 | 0 | 0 | 0 | 0 | 1.506 |
| 1972 | 0 | 1.551 | 0 | 0 | 0 | 0 | 0 | 1.548 |
| 1973 | 2.202 | 1.52 | 0 | 0 | 0 | 0 | 0 | 1.581 |
| 1974 | 1.319 | 1.229 | 0 | 0.833 | 0.89 | 0 | 0 | 1.183 |
| 1975 | 1.19 | 2.523 | 0 | 0 | 0 | 0 | 0 | 1.016 |
| 1976 | 1.243 | 1.806 | 0 | 1.679 | 0 | 0 | 0 | 1.246 |
| 1977 | 1.394 | 1.339 | 1.593 | 0 | 0 | 0 | 0 | 1.325 |
| 1978 | 2.126 | 1.376 | 1.208 | 1.627 | 0 | 0 | 0 | 1.338 |
| 1979 | 1.735 | 1.569 | 1.781 | 1.119 | 1.59 | 0 | 0 | 1.754 |
| 1980 | 1.632 | 1.879 | 2.862 | 0 | 1.482 | 0 | 0 | 1.747 |
| 1981 | 1.366 | 1.314 | 1.785 | 1.587 | 0 | 1.677 | 0 | 1.379 |
| 1982 | 1.738 | 2.068 | 1.543 | 0 | 0 | 0 | 0 | 1.555 |
| 1983 | 1.556 | 1.555 | 1.999 | 0 | 0 | 0 | 0 | 1.572 |
| 1984 | 1.882 | 1.417 | 1.864 | 0 | 0 | 0 | 0 | 1.724 |
| 1985 | 1.695 | 2.014 | 2.152 | 2.741 | 0 | 0 | 4.141 | 1.694 |
| 1986 | 1.517 | 1.832 | 1.925 | 1.504 | 2.635 | 0 | 0 | 1.463 |
| 1987 | 1.149 | 1.851 | 2.774 | 3.04 | 2.828 | 2.664 | 0 | 1.182 |
| 1988 | 0.879 | 1.618 | 0.99 | 3.424 | 3.994 | 4.15 | 0 | 0.984 |
| 1989 | 1.357 | 0.949 | 1.388 | 2.807 | 3.008 | 0 | 0.429 | 1.110 |
| 1990 | 1.682 | 2.288 | 1.964 | 2.506 | 0 | 0 | 0 | 1.860 |
| 1991 | 1.755 | 3.29 | 2.17 | 1.343 | 0 | 0 | 2.869 | 1.201 |
| 1992 | 1.564 | 2.18 | 0 | 0 | 0 | 0 | 0 | 1.639 |
| 1993 | 1.21 | 1.578 | 2.304 | 1.8 | 2.405 | 0 | 0 | 1.483 |
| 1994 | 1.243 | 1.319 | 1.961 | 2.43 | 0 | 0 | 0 | 0.992 |
| 1995 | 0.94 | 1.953 | 1.996 | 2.492 | 0 | 0 | 0 | 1.020 |
| 1996 | 1.163 | 1.046 | 1.141 | 0 | 3.167 | 0 | 0 | 1.137 |
| 1997 | 2.024 | 1.63 | 2.252 | 0 | 3.033 | 0 | 0 | 1.020 |
| 1998 | 1.064 | 2.488 | 2.585 | 3.322 | 2.591 | 0 | 0 | 1.077 |
| 1999 | 1.465 | 1.465 | 3.246 | 1.993 | 2.954 | 2.829 | 0 | 1.172 |
| 2000 | 1.186 | 1.262 | 0 | 2.168 | 0 | 0 | 0 | 0.813 |
| 2001 | 0.975 | 1.089 | 3.361 | 0.597 | 0 | 0 | 0 | 1.015 |
| 2002 | 1.995 | 0.916 | 0 | 2.698 | 0 | 0 | 0 | 0.939 |
| 2003 | 1.232 | 1.393 | 2.682 | 0 | 0 | 0 | 0 | 1.086 |
| 2004 | 0.853 | 1.396 | 3.976 | 0 | 0 | 0 | 0 | 0.988 |
| 2005 | 1.467 | 2.084 | 3.491 | 2.275 | 0 | 0 | 0 | 1.018 |
| 2006 | 1.182 | 1.682 | 2.675 | 0 | 3.889 | 5.471 | 0 | 1.294 |
| 2007 | 0.825 | 2.160 | 2.270 | 0 | 0 | 0 | 0 | 0.685 |
| 2008 | 0.911 | 2.494 | 2.109 | 0 | 0 | 0 | 0 | 0.827 |
| 2009 | 1.042 | 1.233 | 1.874 | 0 | 0 | 0 | 0 | 1.008 |

Table 3.3.6. Haddock in Division VIa. Weights-at-age (kg) in landings. Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 0.000 | 0.273 | 0.295 | 0.440 | 0.695 | 0.916 | 1.041 | 1.249 | 1.517 |
| 1966 | 0.000 | 0.315 | 0.324 | 0.563 | 0.575 | 1.041 | 1.125 | 1.325 | 1.522 |
| 1967 | 0.000 | 0.285 | 0.374 | 0.635 | 0.576 | 0.667 | 1.177 | 1.844 | 1.611 |
| 1968 | 0.000 | 0.259 | 0.367 | 0.627 | 0.827 | 0.731 | 0.811 | 1.430 | 1.903 |
| 1969 | 0.000 | 0.199 | 0.314 | 0.570 | 0.921 | 1.042 | 1.024 | 0.999 | 1.569 |
| 1970 | 0.000 | 0.348 | 0.261 | 0.389 | 0.817 | 1.283 | 1.262 | 1.043 | 1.342 |
| 1971 | 0.000 | 0.295 | 0.328 | 0.360 | 0.549 | 1.040 | 1.313 | 1.651 | 1.426 |
| 1972 | 0.000 | 0.285 | 0.325 | 0.406 | 0.532 | 0.546 | 0.984 | 1.499 | 1.538 |
| 1973 | 0.000 | 0.259 | 0.329 | 0.408 | 0.599 | 0.512 | 0.571 | 1.185 | 1.706 |
| 1974 | 0.000 | 0.264 | 0.328 | 0.393 | 0.530 | 0.764 | 0.685 | 0.798 | 1.142 |
| 1975 | 0.000 | 0.277 | 0.365 | 0.465 | 0.585 | 0.832 | 1.027 | 1.001 | 1.009 |
| 1976 | 0.000 | 0.251 | 0.345 | 0.504 | 0.596 | 0.831 | 1.095 | 1.585 | 1.084 |
| 1977 | 0.000 | 0.307 | 0.370 | 0.437 | 0.689 | 0.800 | 1.128 | 1.337 | 1.117 |
| 1978 | 0.000 | 0.257 | 0.353 | 0.419 | 0.524 | 0.832 | 1.060 | 1.152 | 1.399 |
| 1979 | 0.000 | 0.269 | 0.386 | 0.467 | 0.732 | 0.779 | 1.040 | 1.491 | 1.944 |
| 1980 | 0.000 | 0.251 | 0.373 | 0.587 | 0.722 | 0.998 | 0.985 | 1.143 | 1.565 |
| 1981 | 0.000 | 0.289 | 0.357 | 0.502 | 0.887 | 0.975 | 1.376 | 1.294 | 1.347 |
| 1982 | 0.000 | 0.285 | 0.369 | 0.452 | 0.754 | 1.126 | 1.539 | 1.549 | 1.514 |
| 1983 | 0.000 | 0.479 | 0.424 | 0.518 | 0.568 | 1.004 | 1.370 | 1.716 | 1.558 |
| 1984 | 0.000 | 0.273 | 0.388 | 0.486 | 0.705 | 0.713 | 1.087 | 1.392 | 2.075 |
| 1985 | 0.000 | 0.283 | 0.346 | 0.494 | 0.641 | 0.803 | 0.875 | 1.272 | 1.277 |
| 1986 | 0.000 | 0.294 | 0.373 | 0.440 | 0.637 | 0.903 | 1.115 | 1.043 | 1.418 |
| 1987 | 0.000 | 0.276 | 0.337 | 0.435 | 0.570 | 0.880 | 1.105 | 1.250 | 1.147 |
| 1988 | 0.000 | 0.310 | 0.338 | 0.462 | 0.567 | 0.706 | 1.027 | 1.280 | 1.279 |
| 1989 | 0.000 | 0.372 | 0.406 | 0.468 | 0.625 | 0.749 | 0.894 | 1.115 | 1.462 |
| 1990 | 0.000 | 0.335 | 0.443 | 0.532 | 0.618 | 0.908 | 1.108 | 1.280 | 1.823 |
| 1991 | 0.000 | 0.287 | 0.382 | 0.556 | 0.618 | 0.678 | 0.931 | 1.053 | 1.091 |
| 1992 | 0.000 | 0.310 | 0.384 | 0.461 | 0.777 | 0.892 | 0.932 | 1.407 | 1.493 |
| 1993 | 0.000 | 0.313 | 0.395 | 0.509 | 0.655 | 0.889 | 0.898 | 1.026 | 1.514 |
| 1994 | 0.000 | 0.280 | 0.352 | 0.454 | 0.633 | 0.723 | 0.929 | 0.959 | 0.909 |
| 1995 | 0.000 | 0.293 | 0.375 | 0.415 | 0.567 | 0.833 | 0.978 | 1.322 | 1.059 |
| 1996 | 0.000 | 0.285 | 0.363 | 0.445 | 0.492 | 0.649 | 0.750 | 0.754 | 1.122 |
| 1997 | 0.000 | 0.275 | 0.365 | 0.425 | 0.621 | 0.735 | 0.925 | 1.057 | 0.921 |
| 1998 | 0.000 | 0.265 | 0.331 | 0.416 | 0.524 | 0.689 | 0.802 | 0.951 | 1.006 |
| 1999 | 0.000 | 0.313 | 0.353 | 0.420 | 0.496 | 0.614 | 0.820 | 0.840 | 1.067 |
| 2000 | 0.000 | 0.265 | 0.347 | 0.410 | 0.465 | 0.572 | 0.724 | 0.840 | 0.749 |
| 2001 | 0.000 | 0.243 | 0.332 | 0.457 | 0.439 | 0.538 | 0.657 | 0.808 | 1.029 |
| 2002 | 0.000 | 0.254 | 0.321 | 0.383 | 0.566 | 0.608 | 0.632 | 0.691 | 0.810 |
| 2003 | 0.000 | 0.240 | 0.311 | 0.389 | 0.428 | 0.654 | 0.651 | 0.917 | 0.946 |
| 2004 | 0.000 | 0.253 | 0.329 | 0.394 | 0.391 | 0.448 | 0.541 | 0.718 | 0.782 |
| 2005 | 0.000 | 0.270 | 0.358 | 0.415 | 0.542 | 0.596 | 0.594 | 1.167 | 0.921 |
| 2006 | 0.000 | 0.291 | 0.348 | 0.392 | 0.437 | 0.508 | 0.527 | 0.621 | 1.242 |
| 2007 | 0.000 | 0.248 | 0.357 | 0.398 | 0.423 | 0.458 | 0.558 | 0.605 | 0.682 |
| 2008 | 0.000 | 0.275 | 0.378 | 0.418 | 0.505 | 0.578 | 0.666 | 0.709 | 0.823 |
| 2009 | 0.000 | 0.344 | 0.361 | 0.467 | 0.488 | 0.581 | 0.687 | 0.691 | 0.897 |

Table 3.3.6. Continued. Haddock in Division VIa. Weights-at-age (kg) in landings. Values used in the final assessment are boxed.

| Year | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 1.920 | 1.833 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.713 |
| 1966 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.522 |
| 1967 | 2.355 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.786 |
| 1968 | 2.516 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.005 |
| 1969 | 2.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.590 |
| 1970 | 1.791 | 1.213 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.352 |
| 1971 | 1.466 | 2.042 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.506 |
| 1972 | 0.000 | 1.551 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.548 |
| 1973 | 2.202 | 1.520 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.581 |
| 1974 | 1.319 | 1.229 | 0.000 | 0.833 | 0.890 | 0.000 | 0.000 | 1.183 |
| 1975 | 1.190 | 2.523 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.016 |
| 1976 | 1.243 | 1.806 | 0.000 | 1.679 | 0.000 | 0.000 | 0.000 | 1.246 |
| 1977 | 1.394 | 1.339 | 1.593 | 0.000 | 0.000 | 0.000 | 0.000 | 1.325 |
| 1978 | 2.126 | 1.376 | 1.208 | 1.627 | 0.000 | 0.000 | 0.000 | 1.338 |
| 1979 | 1.735 | 1.569 | 1.781 | 1.119 | 1.590 | 0.000 | 0.000 | 1.754 |
| 1980 | 1.632 | 1.879 | 2.862 | 0.000 | 1.482 | 0.000 | 0.000 | 1.747 |
| 1981 | 1.366 | 1.314 | 1.785 | 1.587 | 0.000 | 1.677 | 0.000 | 1.379 |
| 1982 | 1.738 | 2.068 | 1.543 | 0.000 | 0.000 | 0.000 | 0.000 | 1.555 |
| 1983 | 1.556 | 1.555 | 1.999 | 0.000 | 0.000 | 0.000 | 0.000 | 1.572 |
| 1984 | 1.882 | 1.417 | 1.864 | 0.000 | 0.000 | 0.000 | 0.000 | 1.724 |
| 1985 | 1.695 | 2.014 | 2.152 | 2.741 | 0.000 | 0.000 | 4.141 | 1.694 |
| 1986 | 1.517 | 1.832 | 1.925 | 1.504 | 2.635 | 0.000 | 0.000 | 1.463 |
| 1987 | 1.149 | 1.851 | 2.774 | 3.040 | 2.828 | 2.664 | 0.000 | 1.182 |
| 1988 | 0.879 | 1.618 | 0.990 | 3.424 | 3.994 | 4.150 | 0.000 | 0.984 |
| 1989 | 1.357 | 0.948 | 1.388 | 2.807 | 3.008 | 0.000 | 0.429 | 1.109 |
| 1990 | 1.682 | 2.288 | 1.964 | 2.506 | 0.000 | 0.000 | 0.000 | 1.860 |
| 1991 | 1.755 | 3.290 | 2.170 | 1.343 | 0.000 | 0.000 | 2.869 | 1.201 |
| 1992 | 1.564 | 2.180 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1.639 |
| 1993 | 1.210 | 1.578 | 2.304 | 1.800 | 2.405 | 0.000 | 0.000 | 1.483 |
| 1994 | 1.243 | 1.319 | 1.961 | 2.430 | 0.000 | 0.000 | 0.000 | 0.992 |
| 1995 | 0.940 | 1.953 | 1.996 | 2.492 | 0.000 | 0.000 | 0.000 | 1.020 |
| 1996 | 1.163 | 1.046 | 1.141 | 0.000 | 3.167 | 0.000 | 0.000 | 1.137 |
| 1997 | 2.024 | 1.630 | 2.252 | 0.000 | 3.033 | 0.000 | 0.000 | 1.020 |
| 1998 | 1.064 | 2.488 | 2.585 | 3.322 | 2.591 | 0.000 | 0.000 | 1.077 |
| 1999 | 1.465 | 1.465 | 3.246 | 1.993 | 2.954 | 2.829 | 0.000 | 1.172 |
| 2000 | 1.186 | 1.262 | 0.000 | 2.168 | 0.000 | 0.000 | 0.000 | 0.813 |
| 2001 | 0.975 | 1.089 | 3.361 | 0.597 | 0.000 | 0.000 | 0.000 | 1.015 |
| 2002 | 1.995 | 0.916 | 0.000 | 2.698 | 0.000 | 0.000 | 0.000 | 0.939 |
| 2003 | 1.253 | 1.395 | 2.682 | 0.000 | 0.000 | 0.000 | 0.000 | 1.091 |
| 2004 | 0.853 | 1.396 | 3.976 | 0.000 | 0.000 | 0.000 | 0.000 | 0.988 |
| 2005 | 1.467 | 2.084 | 3.491 | 2.275 | 0.000 | 0.000 | 0.000 | 1.023 |
| 2006 | 1.182 | 1.682 | 2.675 | 0.000 | 3.889 | 5.471 | 0.000 | 1.294 |
| 2007 | 0.825 | 2.160 | 2.270 | 0.000 | 0.000 | 0.000 | 0.000 | 0.685 |
| 2008 | 0.911 | 2.494 | 2.109 | 2.966 | 0.000 | 0.000 | 0.000 | 0.862 |
| 2009 | 1.042 | 1.233 | 1.874 | 0.000 | 3.002 | 0.000 | 0.000 | 1.011 |

Table 3.3.7. Haddock in Division VIa. Weights-at-age (kg) in discards. Values used in the final assessment are boxed.

| Year | AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1965 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1975 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1976 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.040 | 0.156 | 0.215 | 0.265 | 0.279 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1978 | 0.059 | 0.125 | 0.208 | 0.231 | 0.259 | 0.265 | 0.308 | 0.000 | 0.000 |
| 1979 | 0.032 | 0.180 | 0.230 | 0.272 | 0.266 | 0.303 | 0.000 | 0.000 | 0.000 |
| 1980 | 0.077 | 0.120 | 0.243 | 0.287 | 0.334 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.082 | 0.106 | 0.209 | 0.360 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.038 | 0.155 | 0.238 | 0.247 | 0.363 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.050 | 0.165 | 0.237 | 0.283 | 0.298 | 0.536 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.059 | 0.145 | 0.248 | 0.303 | 0.331 | 0.278 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.019 | 0.132 | 0.242 | 0.326 | 0.362 | 0.423 | 0.353 | 0.000 | 0.000 |
| 1986 | 0.064 | 0.173 | 0.193 | 0.248 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.028 | 0.163 | 0.218 | 0.247 | 0.281 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.085 | 0.157 | 0.208 | 0.279 | 0.331 | 0.341 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.052 | 0.193 | 0.226 | 0.237 | 0.491 | 0.961 | 1.423 | 0.000 | 2.572 |
| 1990 | 0.073 | 0.108 | 0.250 | 0.228 | 0.242 | 0.268 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.058 | 0.178 | 0.218 | 0.278 | 0.000 | 0.263 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.050 | 0.130 | 0.247 | 0.258 | 0.242 | 0.000 | 0.947 | 0.000 | 0.000 |
| 1993 | 0.037 | 0.105 | 0.238 | 0.287 | 0.382 | 0.348 | 0.000 | 0.430 | 0.000 |
| 1994 | 0.031 | 0.163 | 0.229 | 0.291 | 0.337 | 0.304 | 0.000 | 0.000 | 0.000 |
| 1995 | 0.030 | 0.144 | 0.243 | 0.281 | 0.310 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.047 | 0.126 | 0.206 | 0.282 | 0.300 | 0.317 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.048 | 0.148 | 0.226 | 0.283 | 0.340 | 0.317 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.089 | 0.151 | 0.251 | 0.298 | 0.337 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.035 | 0.163 | 0.213 | 0.276 | 0.318 | 0.311 | 0.206 | 0.000 | 0.000 |
| 2000 | 0.053 | 0.125 | 0.223 | 0.257 | 0.259 | 0.625 | 0.337 | 0.000 | 0.000 |
| 2001 | 0.050 | 0.109 | 0.211 | 0.243 | 0.254 | 0.245 | 0.000 | 0.000 | 0.000 |
| $2002$ | 0.048 | 0.117 | 0.196 | 0.253 | 0.305 | 0.456 | 0.000 | 0.358 | 0.000 |
| 2003 | 0.036 | 0.123 | 0.223 | 0.233 | 0.282 | 0.462 | 0.439 | 0.496 | 0.591 |
| 2004 | 0.033 | 0.112 | 0.183 | 0.237 | 0.242 | 0.256 | 0.000 | 0.411 | 0.000 |
| 2005 | 0.053 | $0.103$ | 0.190 | 0.262 | 0.320 | 0.290 | 0.322 | 0.416 | 0.493 |
| 2006 | 0.024 | 0.154 | 0.241 | 0.284 | 0.313 | 0.318 | 0.348 | 0.336 | 0.000 |
| 2007 | 0.060 | 0.113 | 0.211 | 0.288 | 0.314 | 0.336 | 0.368 | 0.373 | 0.000 |
| 2008 | 0.022 | 0.112 | 0.226 | 0.287 | 0.322 | 0.389 | 0.312 | 0.458 | 0.419 |
| 2009 | 0.048 | 0.134 | 0.235 | 0.271 | 0.298 | 0.362 | 0.309 | 0.356 | 0.000 |

Table 3.3.7. Continued. Haddock in Division VIa. Weights-at-age (kg) in discards. Values used in the final assessment are boxed.

| Year | AGE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | 8+ |
| 1965 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1966 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1967 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1968 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1969 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1970 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1971 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1972 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1973 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1974 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1975 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1976 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1977 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1978 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1980 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1981 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1982 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1983 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1984 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1985 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1986 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1987 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1988 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1989 | 0.000 | 3.048 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.810 |
| 1990 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $1991$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $1992$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $1994$ | $0.000$ | $0.000$ | $0.000$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $1995$ | $0.000$ | $0.000$ | $0.000$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | $0.000$ | $0.000$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1998 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.432 | 0.689 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.493 |
| 2004 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.493 |
| 2006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.419 |
| 2009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 3.3.8. Haddock in Division VIa. Available research-vessels survey data. Values used in the final assessment are boxed.

| ScoGFS Q1 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Age |  |  |  |  |  |  |  |  |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total |
| 1985 | 1104 | 4085 | 68 | 80 | 141 | 388 | 27 | 1 | 5893 |
| 1986 | 753 | 1669 | 1877 | 17 | 14 | 47 | 90 | 5 | 4467 |
| 1987 | 5518 | 446 | 460 | 690 | 25 | 34 | 25 | 67 | 7198 |
| 1988 | 571 | 3610 | 303 | 112 | 246 | 10 | 4 | 8 | 4856 |
| 1989 | 178 | 488 | 1701 | 98 | 49 | 69 | 5 | 1 | 2588 |
| 1990 | 2577 | 87 | 54 | 296 | 26 | 6 | 36 | 3 | 3082 |
| 1991 | 1591 | 1763 | 92 | 25 | 184 | 9 | 4 | 15 | 3668 |
| 1992 | 3618 | 1193 | 321 | 12 | 13 | 28 | 6 | 1 | 5191 |
| 1993 | 5371 | 5922 | 675 | 167 | 0 | 2 | 18 | 2 | 12155 |
| 1994 | 1151 | 2300 | 787 | 126 | 39 | 3 | 1 | 8 | 4407 |
| 1995 | 7112 | 1074 | 1697 | 485 | 65 | 30 | 10 | 4 | 10473 |
| 1996 | 4401 | 3742 | 315 | 456 | 125 | 20 | 11 | 3 | 9070 |
| 1997 | 4262 | 2018 | 1915 | 147 | 151 | 53 | 2 | 1 | 8548 |
| 1998 | 5034 | 2720 | 616 | 562 | 40 | 64 | 19 | 7 | 9055 |
| 1999 | 941 | 2989 | 687 | 168 | 128 | 15 | 11 | 2 | 4939 |
| 2000 | 7936 | 553 | 440 | 97 | 13 | 20 | 1 | 3 | 9060 |
| 2001 | 3421 | 5762 | 143 | 146 | 34 | 16 | 6 | 1 | 9528 |
| 2002 | 2339 | 3246 | 5293 | 56 | 70 | 24 | 9 | 3 | 11037 |
| 2003 | 2650 | 1696 | 1449 | 1874 | 23 | 34 | 18 | 4 | 7744 |
| 2004 | 1397 | 2765 | 869 | 1199 | 609 | 11 | 3 | 5 | 6853 |
| 2005 | 573 | 633 | 1402 | 351 | 512 | 402 | 5 | 3 | 3878 |
| 2006 | 633 | 892 | 539 | 397 | 156 | 170 | 51 | 2 | 2838 |
| 2007 | 99 | 2019 | 296 | 121 | 192 | 82 | 89 | 65 | 2898 |
| 2008 | 86 | 113 | 1094 | 98 | 84 | 71 | 13 | 15 | 1558 |
| 2009 | 42 | 113 | 147 | 1445 | 29 | 43 | 63 | 7 | 1882 |
| 2010 | 706 | 111 | 26 | 71 | 452 | 23 | 4 | 9 | 1393 |

Table 3.3.8. Continued. Haddock in Division VIa. Available research-vessels survey data. Values used in the final assessment are boxed.


Table 3.3.8. Continued. Haddock in Division VIa. Available research-vessels survey data. Values used in the final assessment are boxed.

| IRGFS |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |  |  |  |  |  |
| Year | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | Total |
| 2003 | 1127 | 207 | 7588 | 2382 | 839 | 355 | 22 | 30 | 7 | 0 | 3 | 2 | 11228 |
| 2004 | 1200 | 86 | 2163 | 3322 | 1281 | 941 | 957 | 60 | 10 | 21 | 0 | 0 | 8755 |
| 2005 | 960 | 233 | 1160 | 767 | 778 | 315 | 87 | 3 | 0 | 0 | 1 | 0 | 3111 |
| 2006 | 1510 | 313 | 207 | 1027 | 381 | 1337 | 543 | 130 | 59 | 0 | 0 | 0 | 3684 |
| 2007 | 1173 | 320 | 979 | 1049 | 346 | 689 | 101 | 64 | 69 | 1 | 0 | 0 | 3298 |
| 2008 | 1135 | 76 | 2052 | 562 | 645 | 74 | 196 | 169 | 31 | 14 | 0 | 0 | 3742 |
| 2009 | 1378 | 744 | 535 | 919 | 309 | 328 | 76 | 187 | 61 | 6 | 0 | 0 | 2422 |

Table 3.3.9. Haddock in Division VIa. TSA parameter estimates from this year's assessment, along with those from previous assessments for comparison. * fixed parameter.

| Parameter | Notation | Description | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Initial fishing mortality | F (1, 1978) | Fishing mortality at age a in year y | 0.42 | 0.28 | 0.26 | 0.23 | 0.25 | 0.40 | 0.40 | 0.43 |
|  | F ( 2,1978 ) |  | 0.67 | 0.5 | 0.51 | 0.50 | 0.56 | 0.71 | 0.70 | 0.81 |
|  | F ( 4,1978 ) |  | 0.53 | 0.51 | 0.51 | 0.51 | 0.52 | 0.56 | 0.57 | 0.59 |
| Survey selectivities | $\Phi(1)$ |  | 3.99 | 2.25 | 2.35 | 2.49 | 2.58 | 2.60 | 2.58 | 3.11 |
| ScoGFS Q1 | $\Phi(2)$ | ScoGFS Q1 survey selectivity at age a | 4.84 | 2.71 | 2.45 | 2.55 | 3.01 | 3.07 | 3.01 | 3.34 |
|  | $\Phi(4)$ |  | 2.1 | 1.51 | 2.11 | 2.19 | 2.04 | 1.92 | 1.94 | 2.24 |
| Survey selectivities | $\Phi(1)$ |  | - | - | - | 1.99 | 1.62 | 1.77 | 1.75 | 2.24 |
| ScoGFS Q4 | $\Phi(2)$ | ScoGFS Q4 survey selectivity at age a | - | - | - | 1.99 | 1.76 | 1.88 | 1.84 | 2.22 |
|  | $\Phi(4)$ |  | - | - | - | 2.25 | 2.39 | 2.61 | 2.64 | 3.44 |
| Fishing mortality standard deviations | $\sigma$ F | Transitory changes in overall F | 0.00 | 0.11 | 0.10 | 0.10 | 0.12 | 0.20 | 0.20 | 0.19 |
|  | $\sigma \mathrm{U}$ | Persistent changes in selection (age effect in F) | 0.05 | 0.04 | 0.01 | 0.00 | 0.09 | 0.03 | 0.03 | 0.05 |
|  | $\sigma \mathrm{V}$ | Transitory changes in the year effect in F | 0.27 | 0.23 | 0.22 | 0.23 | 0.23 | 0.33 | 0.35 | 0.26 |
|  | $\sigma Y$ | Persistent changes in the year effect in F | 0.00 | 0.14 | 0.09 | 0.09 | 0.07 | 0.00 | 0.00 | 0.15 |
| Survey catchability standard deviations | $\sigma \Omega 1$ | Transitory changes in ScoGFS Q1 catchability | 0.00 | 0.08 | 0.18 | 0.30 | 0.19 | 0.12 | 0.12 | 0.27 |
|  | $\sigma \beta 1$ | Persistent changes in ScoGFS Q1 catchability | 0.14 | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* |
|  | $\sigma \Omega 2$ | Transitory changes in ScoGFS Q4 catchability | - | - | - |  | 0.16 | 0.20 | 0.19 | 0.21 |
|  | $\sigma \beta 2$ | Persistent changes in ScoGFS Q4 catchability | - | - | - |  | 0.00* | 0.00* | 0.00* | 0.00* |
| Measurement coefficients of variation | cV landings | Coefficent of variation of landings-at-age data | 0.22 | 0.25 | 0.23 | 0.20 | 0.20 | 0.24 | 0.25 | 0.28 |
|  | cv discards | Coefficent of variation of discards-at-age data | 0.51 | 0.43 | 0.45 | 0.42 | 0.41 | 0.54 | 0.54 | 0.59 |
|  | cv survey | Coefficent of variation of ScoGFS Q1 survey data | 0.40 | 0.34 | 0.53 | 0.57 | 0.33 | 0.35 | 0.36 | 0.41 |
|  | cv survey | Coefficent of variation of ScoGFS Q4 survey data | - | - | - | 0.57 | 0.22 | 0.34 | 0.35 | 0.51 |

Table 3.3.9. Continued. Haddock in Division VIa. TSA parameter estimates from this year's assessment, along with those from previous assessments for comparison. ${ }^{*}=$ fixed parameter.

| Parameter | Notation | Description | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Discard curve parameters | $\sigma \mathrm{P}$ | Transitory changes in overall discard proportion | 0.50 | 0.19 | 0.20 | 0.19 | 0.18 | 0.20 | 0.20 | 0.00 |
|  | $\sigma \alpha 1$ | Transitory changes in discard-ogive intercept | 0.00 | 0.15 | 0.02 | 0.00 | 0.14 | 0.00 | 0.00 | 0.01 |
|  | бv1 | Persistent changes in discard-ogive intercept | 0.26 | 0.21 | 0.22 | 0.21 | 0.32 | 0.26 | 0.25 | 0.29 |
|  | $\sigma$ o2 | Transitory changes in discard-ogive slope | 0.34 | 0.01 | 0.03 | 0.21 | 0.23 | 0.22 | 0.23 | 0.40 |
|  | $\sigma \vee 2$ | Persistent changes in discard-ogive slope | 0.02 | 0.61 | 0.43 | 0.23 | 0.002 | 0.000 | 0.000 | 0.00 |
| Trend parameters | $\theta \vee 1$ | Trend parameter for discard-ogive intercept | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* |
|  | $\theta \vee 2$ | Trend parameter for discard-ogive slope | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* | 0.00* |
| Recruitment | $\eta 1$ | Ricker parameter (slope at the origin) | 9.10 | 9.63 | 9.71 | 9.73 | 9.06 | 11.35 | 11.08 | 9.62 |
|  | $\eta 2$ | Ricker parameter (curve dome occurs at $1 / \mathfrak{2} 2$ ) | 0.33 | 0.29 | 0.31 | 0.29 | 0.30 | 0.35 | 0.35 | 0.39 |
|  | cv rec | Coefficent of variation of recruitment curve | 0.52 | 0.89 | 0.89 | 0.90 | 0.62 | 0.60 | 0.61 | 0.69 |

Table 3.3.10. Haddock in Division VIa. Estimates of population abundance (in thousands) from the final TSA run.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1978 | 70152 | 7863 | 2533 | 59405 | 4443 | 626 | 463 | 1016 |
| 1979 | 148478 | 42687 | 3868 | 1113 | 22687 | 1500 | 232 | 561 |
| 1980 | 451131 | 84749 | 17041 | 1560 | 417 | 7555 | 465 | 255 |
| 1981 | 63077 | 300652 | 43940 | 6859 | 588 | 183 | 2976 | 282 |
| 1982 | 71998 | 43048 | 181310 | 22186 | 3427 | 295 | 96 | 1621 |
| 1983 | 44841 | 49441 | 25964 | 100081 | 11362 | 1793 | 151 | 907 |
| 1984 | 307976 | 27502 | 26548 | 11833 | 45177 | 5090 | 782 | 472 |
| 1985 | 72879 | 189959 | 11804 | 9940 | 4875 | 19744 | 2079 | 510 |
| 1986 | 58508 | 41641 | 95027 | 5161 | 4071 | 2201 | 7919 | 1108 |
| 1987 | 245833 | 38991 | 23105 | 48687 | 2556 | 2070 | 1166 | 4561 |
| 1988 | 20765 | 138867 | 14681 | 8171 | 16865 | 821 | 647 | 1965 |
| 1989 | 18186 | 10465 | 60424 | 5381 | 2838 | 5869 | 301 | 935 |
| 1990 | 93427 | 9224 | 4257 | 23830 | 1931 | 929 | 1907 | 404 |
| 1991 | 124036 | 57775 | 3423 | 1786 | 9692 | 779 | 385 | 934 |
| 1992 | 167262 | 69017 | 23743 | 1272 | 689 | 3497 | 292 | 483 |
| 1993 | 161499 | 109314 | 33574 | 9895 | 527 | 313 | 1471 | 335 |
| 1994 | 56927 | 99162 | 41312 | 9985 | 2990 | 147 | 81 | 519 |
| 1995 | 184980 | 30799 | 48872 | 16918 | 3750 | 1139 | 60 | 231 |
| 1996 | 100407 | 112926 | 14615 | 21685 | 6502 | 1567 | 441 | 116 |
| 1997 | 119655 | 56585 | 50732 | 5890 | 8117 | 2249 | 626 | 196 |
| 1998 | 129373 | 66750 | 22666 | 17932 | 1926 | 2786 | 636 | 238 |
| 1999 | 27416 | 72524 | 27916 | 8532 | 4924 | 695 | 1071 | 253 |
| 2000 | 456691 | 15879 | 31645 | 11404 | 3088 | 1332 | 273 | 512 |
| 2001 | 166094 | 251386 | 5983 | 10380 | 3266 | 937 | 258 | 211 |
| 2002 | 95068 | 109918 | 132767 | 2731 | 3806 | 1261 | 389 | 157 |
| 2003 | 108188 | 65557 | 68786 | 79895 | 1311 | 1884 | 718 | 271 |
| 2004 | 39439 | 66424 | 32480 | 35010 | 26527 | 377 | 610 | 237 |
| 2005 | 29217 | 25094 | 35577 | 16913 | 16252 | 10706 | 127 | 328 |
| 2006 | 84889 | 17602 | 12459 | 15186 | 6326 | 6428 | 3049 | 146 |
| 2007 | 19210 | 53588 | 8863 | 6384 | 6162 | 2617 | 2762 | 1099 |
| 2008 | 7769 | 12209 | 33781 | 4591 | 3073 | 2699 | 1290 | 1687 |
| 2009 | 7902 | 5152 | 7439 | 21374 | 2377 | 1677 | 1379 | 1545 |
|  |  |  |  |  |  |  |  |  |
| 2010* | 41994 | 5447 | 3269 | 4980 | 12417 | 1341 | 963 | 1603 |
| 2011* | 76211 | 28875 | 3363 | 1995 | 2562 | 6477 | 691 | 1322 |

*Estimates for 2010 and 2011 are TSA forecasts.

Table 3.3.11. Haddock in Division VIa. Standard errors of estimates of population abundance (in thousands) from the final TSA run.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1978 | 8678 | 740 | 313 | 409 | 1212 | 225 | 146 | 345 |
| 1979 | 16206 | 4471 | 313 | 146 | 2440 | 604 | 126 | 202 |
| 1980 | 42779 | 9038 | 2489 | 189 | 79 | 1411 | 313 | 128 |
| 1981 | 5753 | 27493 | 5581 | 1207 | 106 | 47 | 794 | 180 |
| 1982 | 8459 | 3715 | 17299 | 2845 | 587 | 59 | 27 | 441 |
| 1983 | 6750 | 5621 | 2139 | 9133 | 1356 | 298 | 34 | 226 |
| 1984 | 43293 | 3658 | 2822 | 754 | 3392 | 505 | 119 | 85 |
| 1985 | 8532 | 21270 | 1652 | 1426 | 545 | 2220 | 370 | 91 |
| 1986 | 6348 | 4463 | 10668 | 641 | 544 | 321 | 1401 | 227 |
| 1987 | 35798 | 3862 | 2702 | 5433 | 286 | 261 | 190 | 778 |
| 1988 | 4646 | 16935 | 1574 | 1074 | 2070 | 120 | 147 | 383 |
| 1989 | 4420 | 1654 | 7306 | 667 | 406 | 881 | 62 | 201 |
| 1990 | 11828 | 1857 | 638 | 3238 | 274 | 182 | 452 | 114 |
| 1991 | 13891 | 7065 | 593 | 225 | 1173 | 108 | 79 | 201 |
| 1992 | 17194 | 6591 | 2969 | 198 | 80 | 514 | 54 | 99 |
| 1993 | 18958 | 11191 | 3015 | 1180 | 64 | 36 | 235 | 54 |
| 1994 | 12901 | 12954 | 4426 | 1044 | 341 | 14 | 14 | 86 |
| 1995 | 28169 | 7453 | 8225 | 3053 | 657 | 217 | 13 | 56 |
| 1996 | 21886 | 19664 | 3442 | 4077 | 1368 | 265 | 94 | 32 |
| 1997 | 23382 | 12263 | 9887 | 1163 | 1453 | 437 | 89 | 36 |
| 1998 | 23243 | 12732 | 4755 | 3596 | 339 | 377 | 112 | 42 |
| 1999 | 9858 | 13446 | 5276 | 1503 | 1127 | 95 | 137 | 42 |
| 2000 | 106685 | 5383 | 6771 | 2152 | 566 | 398 | 33 | 71 |
| 2001 | 25574 | 50684 | 1777 | 2083 | 566 | 147 | 69 | 49 |
| 2002 | 17011 | 14298 | 21528 | 463 | 528 | 130 | 44 | 26 |
| 2003 | 16480 | 10991 | 8676 | 11966 | 212 | 220 | 52 | 31 |
| 2004 | 7314 | 10410 | 5332 | 4888 | 4473 | 79 | 97 | 50 |
| 2005 | 4525 | 4033 | 5728 | 2617 | 2220 | 1775 | 20 | 46 |
| 2006 | 7918 | 2077 | 1280 | 1850 | 622 | 597 | 364 | 24 |
| 2007 | 3331 | 4655 | 1287 | 673 | 727 | 252 | 251 | 128 |
| 2008 | 3277 | 1907 | 3600 | 724 | 314 | 324 | 118 | 189 |
| 2009 | 9414 | 2229 | 1220 | 2530 | 441 | 191 | 165 | 191 |
|  |  |  |  |  |  |  |  |  |
| 2010* | 31715 | 6795 | 1576 | 937 | 1873 | 299 | 125 | 230 |
| 2011* | 52696 | 21983 | 4226 | 1003 | 695 | 1595 | 209 | 316 |

*Estimates for 2009 and 2010 are TSA forecasts.

Table 3.3.12. Haddock in Division VIa. Estimates of fishing mortality from the final TSA run.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1978 | 0.269 | 0.350 | 0.610 | 0.771 | 0.774 | 0.741 | 0.717 | 0.730 |
| 1979 | 0.359 | 0.694 | 0.703 | 0.787 | 0.863 | 0.826 | 0.837 | 0.835 |
| 1980 | 0.199 | 0.463 | 0.630 | 0.750 | 0.611 | 0.686 | 0.674 | 0.656 |
| $1981$ | $0.184$ | 0.295 | 0.485 | 0.490 | 0.490 | 0.443 | 0.500 | 0.483 |
| 1982 | 0.159 | 0.270 | 0.365 | 0.469 | 0.448 | 0.465 | 0.477 | 0.449 |
| 1983 | 0.315 | 0.419 | 0.311 | 0.407 | 0.448 | 0.464 | 0.458 | 0.503 |
| 1984 | 0.283 | 0.606 | 0.753 | 0.684 | 0.612 | 0.690 | 0.703 | 0.672 |
| 1985 | 0.354 | 0.490 | 0.623 | 0.686 | 0.595 | 0.714 | 0.652 | 0.633 |
| 1986 | 0.182 | 0.390 | 0.461 | 0.471 | 0.445 | 0.417 | 0.464 | 0.470 |
| 1987 | 0.372 | 0.758 | 0.839 | 0.854 | 0.926 | 0.963 | 0.885 | 0.857 |
| 1988 | 0.436 | 0.632 | 0.803 | 0.857 | 0.855 | 0.799 | 0.807 | 0.831 |
| $1989$ | $0.431$ | 0.654 | 0.726 | 0.819 | 0.895 | 0.914 | 0.913 | 0.903 |
| 1990 | 0.282 | 0.755 | 0.664 | 0.685 | 0.691 | 0.654 | 0.689 | 0.690 |
| 1991 | 0.353 | 0.689 | 0.777 | 0.736 | 0.818 | 0.767 | 0.828 | 0.779 |
| 1992 | 0.198 | 0.447 | 0.657 | 0.661 | 0.542 | 0.619 | 0.604 | 0.582 |
| 1993 | 0.286 | 0.708 | 0.932 | 0.936 | 0.860 | 1.005 | 0.941 | 0.964 |
| 1994 | 0.417 | 0.491 | 0.681 | 0.768 | 0.759 | 0.692 | 0.780 | 0.746 |
| 1995 | 0.294 | 0.545 | 0.612 | 0.755 | 0.670 | 0.747 | 0.720 | 0.724 |
| 1996 | 0.372 | 0.600 | 0.700 | 0.786 | 0.860 | 0.718 | 0.846 | 0.809 |
| 1997 | 0.388 | 0.720 | 0.840 | 0.895 | 0.822 | 1.063 | 1.070 | 0.959 |
| 1998 | 0.381 | 0.674 | 0.773 | 1.089 | 0.793 | 0.757 | 1.083 | 0.922 |
| 1999 | 0.342 | 0.626 | 0.695 | 0.833 | 1.018 | 0.736 | 0.725 | 0.844 |
| 2000 | 0.390 | 0.802 | 0.909 | 1.047 | 0.975 | 1.418 | 1.101 | 1.119 |
| 2001 | 0.211 | 0.448 | 0.635 | 0.777 | 0.643 | 0.678 | 0.996 | 0.764 |
| 2002 | 0.170 | 0.266 | 0.319 | 0.538 | 0.504 | 0.363 | 0.509 | 0.483 |
| 2003 | 0.292 | 0.506 | 0.451 | 0.874 | 1.045 | 0.926 | 1.369 | 0.928 |
| 2004 | 0.253 | 0.426 | 0.453 | 0.551 | 0.708 | 0.889 | 0.759 | 0.720 |
| 2005 | 0.311 | 0.498 | 0.652 | 0.777 | 0.724 | 1.057 | 1.067 | 0.894 |
| 2006 | 0.236 | 0.473 | 0.467 | 0.696 | 0.682 | 0.644 | 0.874 | 0.719 |
| 2007 | 0.251 | 0.261 | 0.443 | 0.531 | 0.613 | 0.505 | 0.643 | 0.579 |
| 2008 | 0.186 | 0.294 | 0.256 | 0.452 | 0.405 | 0.467 | 0.462 | 0.447 |
| 2009 | 0.154 | 0.244 | 0.201 | 0.342 | 0.373 | 0.355 | 0.447 | 0.362 |
|  |  |  |  |  |  |  |  |  |
| 2010* | 0.175 | 0.282 | 0.294 | 0.465 | 0.451 | 0.463 | 0.464 | 0.462 |
| 2011* | 0.175 | 0.285 | 0.295 | 0.467 | 0.467 | 0.467 | 0.467 | 0.467 |

*Estimates for 2009 and 2010 are TSA forecasts.

Table 3.3.13. Haddock in Division VIa. Standard errors of estimates of log fishing mortality from the final TSA run.

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| 1978 | 0.266 | 0.180 | 0.180 | 0.153 | 0.174 | 0.204 | 0.218 | 0.212 |
| 1979 | 0.234 | 0.167 | 0.150 | 0.169 | 0.158 | 0.188 | 0.216 | 0.212 |
| 1980 | 0.293 | 0.183 | 0.184 | 0.162 | 0.191 | 0.179 | 0.217 | 0.220 |
| 1981 | 0.291 | 0.199 | 0.176 | 0.182 | 0.191 | 0.214 | 0.213 | 0.228 |
| 1982 | 0.278 | 0.188 | 0.172 | 0.171 | 0.183 | 0.204 | 0.229 | 0.211 |
| 1983 | 0.232 | 0.163 | 0.172 | 0.157 | 0.171 | 0.189 | 0.218 | 0.206 |
| 1984 | 0.368 | 0.167 | 0.153 | 0.142 | 0.154 | 0.193 | 0.214 | 0.218 |
| 1985 | 0.234 | 0.172 | 0.173 | 0.160 | 0.172 | 0.179 | 0.213 | 0.218 |
| 1986 | 0.277 | 0.180 | 0.174 | 0.177 | 0.183 | 0.197 | 0.213 | 0.222 |
| 1987 | 0.261 | 0.138 | 0.149 | 0.138 | 0.143 | 0.179 | 0.209 | 0.196 |
| 1988 | 0.270 | 0.165 | 0.146 | 0.148 | 0.151 | 0.185 | 0.212 | 0.204 |
| 1989 | 0.282 | 0.180 | 0.164 | 0.151 | 0.158 | 0.172 | 0.214 | 0.209 |
| 1990 | 0.263 | 0.166 | 0.182 | 0.166 | 0.170 | 0.189 | 0.208 | 0.218 |
| 1991 | 0.244 | 0.163 | 0.174 | 0.154 | 0.155 | 0.185 | 0.213 | 0.206 |
| 1992 | 0.270 | 0.169 | 0.167 | 0.167 | 0.170 | 0.186 | 0.216 | 0.214 |
| 1993 | 0.264 | 0.146 | 0.126 | 0.134 | 0.136 | 0.185 | 0.196 | 0.213 |
| 1994 | 0.299 | 0.234 | 0.218 | 0.195 | 0.203 | 0.239 | 0.252 | 0.250 |
| 1995 | 0.451 | 0.319 | 0.297 | 0.278 | 0.280 | 0.278 | 0.301 | 0.301 |
| 1996 | 0.443 | 0.298 | 0.301 | 0.276 | 0.271 | 0.273 | 0.272 | 0.295 |
| 1997 | 0.419 | 0.275 | 0.257 | 0.244 | 0.238 | 0.240 | 0.213 | 0.270 |
| 1998 | 0.425 | 0.277 | 0.268 | 0.238 | 0.242 | 0.243 | 0.227 | 0.273 |
| 1999 | 0.455 | 0.290 | 0.281 | 0.262 | 0.255 | 0.251 | 0.213 | 0.278 |
| 2000 | 0.439 | 0.282 | 0.253 | 0.242 | 0.242 | 0.241 | 0.196 | 0.268 |
| 2001 | 0.439 | 0.292 | 0.274 | 0.251 | 0.257 | 0.250 | 0.240 | 0.279 |
| 2002 | 0.463 | 0.303 | 0.298 | 0.273 | 0.268 | 0.258 | 0.214 | 0.283 |
| 2003 | 0.424 | 0.276 | 0.269 | 0.229 | 0.218 | 0.218 | 0.193 | 0.261 |
| 2004 | 0.469 | 0.300 | 0.295 | 0.266 | 0.260 | 0.256 | 0.238 | 0.285 |
| 2005 | 0.446 | 0.270 | 0.225 | 0.183 | 0.182 | 0.186 | 0.192 | 0.234 |
| 2006 | 0.302 | 0.192 | 0.183 | 0.159 | 0.149 | 0.150 | 0.164 | 0.216 |
| 2007 | 0.311 | 0.217 | 0.204 | 0.169 | 0.165 | 0.160 | 0.158 | 0.218 |
| 2008 | 0.317 | 0.229 | 0.236 | 0.192 | 0.184 | 0.172 | 0.173 | 0.225 |
| 2009 | 0.332 | 0.263 | 0.264 | 0.214 | 0.215 | 0.191 | 0.210 | 0.240 |
|  |  |  |  |  |  |  |  |  |
| 2010* | 0.558 | 0.431 | 0.429 | 0.405 | 0.404 | 0.404 | 0.405 | 0.405 |
| 2011* | 0.582 | 0.463 | 0.461 | 0.438 | 0.438 | 0.438 | 0.438 | 0.438 |

*Estimates for 2009 and 2010 are TSA forecasts.

Table 3.3.14. Haddock in Division VIa. Stock summary from final TSA run. "Obs." denotes the SOP of numbers and mean weights-at-age, rather than the reported caught, landed and discarded yield. "Pred." are TSA estimates, and "SE" denotes standard errors. *Estimates for 2010 and 2011 are TSA projections.

| Year | LANDINGS (tonnes) |  |  | DISCARDS (TONNES) |  |  | Total catches (tonnes) |  |  | Mean F(2-6) |  | SSB (TONNES) |  | TSB (TONNES) |  | Recruitment (000s at age 1) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE | Obs. | Pred. | SE | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1978 | 17178 | 18986 | 1836 | 2327 | 2227 | 569 | 19505 | 21164 | 1983 | 0.65 | 0.07 | 39101 | 1183 | 49408 | 1675 | 70152 | 8678 |
| 1979 | 14820 | 15719 | 1844 | 13857 | 9679 | 2155 | 28678 | 25890 | 3078 | 0.77 | 0.07 | 30974 | 2355 | 64042 | 4155 | 148478 | 16206 |
| 1980 | 12759 | 13352 | 1739 | 4715 | 12994 | 3130 | 17474 | 27828 | 4253 | 0.63 | 0.07 | 35115 | 2820 | 107401 | 7015 | 451131 | 42779 |
| 1981 | 18233 | 18262 | 2659 | 15048 | 11990 | 2662 | 33281 | 30749 | 4461 | 0.44 | 0.05 | 73728 | 4997 | 121309 | 7705 | 63077 | 5753 |
| 1982 | 29635 | 26511 | 4057 | 10063 | 5979 | 1301 | 39698 | 31005 | 4161 | 0.40 | 0.05 | 98368 | 7050 | 114733 | 7193 | 71998 | 8459 |
| 1983 | 29405 | 26348 | 3378 | 6787 | 5211 | 1034 | 36192 | 31423 | 3611 | 0.41 | 0.04 | 91221 | 5613 | 105174 | 5871 | 44841 | 6750 |
| 1984 | 30012 | 28127 | 2537 | 16343 | 12203 | 4140 | 46355 | 39554 | 5366 | 0.67 | 0.06 | 62780 | 2983 | 112542 | 7349 | 307976 | 43293 |
| 1985 | 24393 | 24454 | 2664 | 17444 | 13435 | 2886 | 41837 | 37188 | 4603 | 0.62 | 0.06 | 65580 | 4342 | 97512 | 6724 | 72879 | 8532 |
| 1986 | 19561 | 20039 | 2809 | 7153 | 4246 | 895 | 26714 | 22951 | 3002 | 0.44 | 0.05 | 60041 | 4577 | 75525 | 4950 | 58508 | 6348 |
| 1987 | 27012 | 29160 | 3015 | 16193 | 13486 | 3725 | 43205 | 42635 | 5163 | 0.87 | 0.07 | 54615 | 3814 | 100544 | 7473 | 245833 | 35798 |
| 1988 | 21136 | 21785 | 2466 | 9536 | 8684 | 2026 | 30672 | 30137 | 3806 | 0.79 | 0.07 | 46703 | 3311 | 65402 | 5094 | 20765 | 4646 |
| 1989 | 16688 | 18552 | 2596 | 2981 | 2826 | 735 | 19669 | 20734 | 2784 | 0.80 | 0.08 | 38211 | 3304 | 43675 | 3577 | 18186 | 4420 |
| 1990 | 10135 | 11276 | 1689 | 5387 | 2944 | 739 | 15522 | 13274 | 1837 | 0.69 | 0.07 | 22435 | 2092 | 34383 | 2655 | 93427 | 11828 |
| 1991 | 10557 | 10064 | 1138 | 8691 | 9355 | 2041 | 19248 | 19983 | 2723 | 0.76 | 0.07 | 21574 | 1628 | 51813 | 3778 | 124036 | 13891 |
| 1992 | 11350 | 9482 | 1185 | 9163 | 7798 | 1446 | 20513 | 18012 | 2197 | 0.59 | 0.06 | 29046 | 2019 | 60707 | 3774 | 167262 | 17194 |
| 1993 | 19060 | 17672 | 1814 | 16811 | 14156 | 2266 | 35871 | 31760 | 3088 | 0.89 | 0.08 | 41789 | 2628 | 72328 | 4592 | 161499 | 18958 |
| 1994 | 14243 | 12069 | 1650 | 11098 | 11492 | 2388 | 25342 | 23827 | 3123 | 0.68 | 0.11 | 39929 | 3181 | 60337 | 5342 | 56927 | 12901 |
| 1995 | 12368 | 13458 | 3880 | 8552 | 10319 | 3553 | 20920 | 23481 | 6342 | 0.67 | 0.16 | 35986 | 5067 | 67096 | 7911 | 184980 | 28169 |

Continued on next page.

Table 3.3.14. Continued. Haddock in Division VIa. Stock summary from final TSA run. "Obs." denotes the SOP of numbers and mean weights-at-age, rather than the reported caught, landed and discarded yield. "Pred." are TSA estimates, and "SE" denotes standard errors. *Estimates for 2010 and 2011 are TSA projections.

| Year | LANDINGS (tonnes) |  |  | DISCARDS (TONNES) |  |  | Total catches (tonnes) |  |  | Mean F(2-6) |  | SSB (TONNES) |  | TSB (TONNES) |  | Recruitment (000s at age 1) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Obs. | Pred. | SE | Obs. | Pred. | SE | Obs. | Pred. | SE | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1996 | 13453 | 13322 | 3998 | 11364 | 12192 | 3775 | 24817 | 25742 | 6920 | 0.73 | 0.17 | 37416 | 5295 | 62091 | 8232 | 100407 | 21886 |
| 1997 | 12874 | 15680 | 4406 | 6470 | 12906 | 3914 | 19344 | 29612 | 7009 | 0.87 | 0.17 | 40998 | 5911 | 65680 | 8351 | 119655 | 23382 |
| 1998 | 14401 | 12214 | 3421 | 5535 | 13435 | 4069 | 19936 | 26674 | 6579 | 0.82 | 0.16 | 32835 | 4620 | 62221 | 7731 | 129373 | 23243 |
| 1999 | 10430 | 9438 | 2869 | 4891 | 8359 | 2630 | 15321 | 18598 | 4716 | 0.78 | 0.17 | 29680 | 4074 | 42351 | 5918 | 27416 | 9858 |
| 2000 | 6952 | 10023 | 2973 | 7899 | 20164 | 8914 | 14851 | 30681 | 10305 | 1.03 | 0.20 | 22344 | 3650 | 82240 | 14996 | 456691 | 106685 |
| 2001 | 6731 | 7468 | 2843 | 6657 | 18812 | 6369 | 13389 | 27749 | 8695 | 0.64 | 0.13 | 44261 | 7767 | 89001 | 13798 | 166094 | 25574 |
| 2002 | 7097 | 9155 | 3615 | 8880 | 9695 | 3090 | 15977 | 18365 | 4969 | 0.40 | 0.09 | 58826 | 7431 | 79937 | 8532 | 95068 | 17011 |
| 2003 | 5334 | 23301 | 5907 | 4104 | 11820 | 3432 | 9438 | 33355 | 6388 | 0.76 | 0.14 | 61898 | 6163 | 81986 | 7094 | 108188 | 16480 |
| 2004 | 3199 | 13306 | 3641 | 4380 | 6125 | 1896 | 7579 | 17267 | 4243 | 0.61 | 0.13 | 38256 | 4320 | 48062 | 5116 | 39439 | 7314 |
| 2005 | 3148 | 15803 | 3427 | 3546 | 4911 | 1462 | 6694 | 17847 | 3477 | 0.74 | 0.11 | 34023 | 3702 | 39171 | 3954 | 29217 | 4525 |
| 2006 | 5723 | 7178 | 958 | 5161 | 5109 | 1019 | 10884 | 11885 | 1415 | 0.59 | 0.06 | 20580 | 1263 | 35621 | 1978 | 84889 | 7918 |
| 2007 | 3735 | 4315 | 515 | 4009 | 3279 | 629 | 7745 | 7544 | 922 | 0.47 | 0.05 | 18834 | 1131 | 26096 | 1590 | 19210 | 3331 |
| 2008 | 2792 | 3885 | 489 | 1285 | 1960 | 495 | 4077 | 5998 | 869 | 0.37 | 0.05 | 22114 | 1729 | 24283 | 1936 | 7769 | 3277 |
| 2009 | 2709 | 3629 | 568 | 1676 | 1036 | 359 | 4385 | 4487 | 628 | 0.30 | 0.05 | 16818 | 1615 | 18430 | 2368 | 7902 | 9414 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2010* | NA | 4026 | 1262 | NA | 1182 | 832 | NA | 5140 | 1770 | 0.39 | 0.14 | 13377 | 2156 | 19018 | 5131 | 41994 | 31715 |
| 2011* | NA | 3023 | 1002 | NA | 2829 | 1775 | NA | 5774 | 2463 | 0.40 | 0.16 | 12726 | 4374 | 24904 | 9259 | 76211 | 52696 |
| Min | 2709 | 3629 |  | 1285 | 1036 |  | 4077 | 4487 |  | 0.30 | 0.00 | 16818 |  | 18430 |  | 7769 |  |
| GM | 11007 | 13292 |  | 6695 | 7410 |  | 18293 | 21610 |  | 0.63 | 0.00 | 38575 |  | 61218 |  | 75905 |  |
| AM | 13660 | 15126 |  | 8063 | 9026 |  | 21723 | 23981 |  | 0.65 | 0.00 | 42690 |  | 67534 |  | 117290 |  |
| Max | 30012 | 29160 |  | 17444 | 20164 |  | 46355 | 42635 |  | 1.03 | 0.00 | 98368 |  | 121309 |  | 456691 |  |

Table 3.3.15. Haddock in Division VIa. Mean weights-at-age in total catches (or stock) and forecasted weights-at-age in 2008. Forecasts in this table are based on either of simple three year means or linear model projections: those that were used in the forecasts are shaded and boxed: simple three year means were used for the younger ages (1-2) and linear model projections for the older ages (3-8+). The weights for the 1999 year-class are highlighted in red.


Table 3.3.16. Haddock in Division VIa. Inputs to short-term forecasts.

| Label | Value | CV | Label | Value | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Number-at-age |  |  | Stock weight |  |  |
| N1 | 41994 | 0.76 | WS1 | 0.121 | 0.10 |
| N2 | 5447 | 1.25 | WS2 | 0.239 | 0.07 |
| N3 | 3269 | 0.48 | WS3 | 0.352 | 0.05 |
| N4 | 4980 | 0.19 | WS4 | 0.435 | 0.10 |
| N5 | 12417 | 0.15 | WS5 | 0.532 | 0.13 |
| N6 | 1341 | 0.22 | WS6 | 0.600 | 0.06 |
| N7 | 963 | 0.13 | WS7 | 0.644 | 0.08 |
| N8 | 1603 | 0.14 | WS8 | 0.749 | 0.22 |
| Removals selectivity |  |  | Removals weights |  |  |
| sH1 | 0.197 | 0.17 | WH1 | 0.121 | 0.10 |
| sH2 | 0.266 | 0.43 | WH2 | 0.239 | 0.07 |
| sH3 | 0.300 | 0.39 | WH3 | 0.352 | 0.05 |
| sH4 | 0.442 | 0.28 | WH4 | 0.435 | 0.10 |
| sH5 | 0.464 | 0.31 | WH5 | 0.532 | 0.13 |
| sH6 | 0.442 | 0.21 | WH6 | 0.600 | 0.06 |
| sH7 | 0.517 | 0.40 | WH7 | 0.644 | 0.08 |
| sH8 | 0.463 | 0.29 | WH8 | 0.749 | 0.22 |
| Natural mortality |  |  | Prop.mature. |  |  |
| M1 | 0.2 | 0.1 | MT1 | 0 | 0.1 |
| M2 | 0.2 | 0.1 | MT2 | 0.57 | 0.1 |
| M3 | 0.2 | 0.1 | MT3 | 1 | 0.1 |
| M4 | 0.2 | 0.1 | MT4 | 1 | 0 |
| M5 | 0.2 | 0.1 | MT5 | 1 | 0 |
| M6 | 0.2 | 0.1 | MT6 | 1 | 0 |
| M7 | 0.2 | 0.1 | MT7 | 1 | 0 |
| M8 | 0.2 | 0.1 | MT8 | 1 | 0 |
| Relative effort |  |  | Year effect for M |  |  |
| 'HF09' | 1 | 0.08 | 'K09' | 1 | 0.1 |
| 'HF10' | 1 | 0.08 | 'K10' | 1 | 0.1 |
| 'HF11' | 1 | 0.08 | 'K11' | 1 | 0.1 |
| Recruitment |  |  |  |  |  |
| 'R11' | 76211 | 0.39 |  |  |  |
| 'R12' | 75905 | 1.48 |  |  |  |
| Prop. F before spawning | 0 |  |  |  |  |
| Prop. M before spawning | 0 |  |  |  |  |

Stock numbers in 2010 are TSA survivors.

Table 3.3.17. Haddock in Division VIa. Catch forecast output and estimates of coefficient of variation (CV) from linear analysis. Catch included 1978-1994 and 2006-2009.



Figure 3.3.1. Haddock in Division VIa. Mean weights-at-age (kg) in total catch (also used for stock weights). Dotted lines show loess smoothers fitted through each time-series at age. For clarity, only ages 1-8+ are shown here.

Landings


Figure 3.3.2. Haddock in Division VIa. Mean weights-at-age (kg) in landings. Dotted lines show Loess smoothers fitted through each time-series at age. For clarity, only ages 1-8+ are shown here.

## Discards



Figure 3.3.3. Haddock in Division VIa. Mean weights-at-age (kg) in discards. Dotted lines show Loess smoothers fitted through each time-series at age. For clarity, only ages 1-4 are shown here.


Figure 3.3.4. Haddock in Division VIa. TSA stock summaries from the final run with catch data included 1978-1994 and 2006-2009. Estimates are plotted with approximate pointwise $\mathbf{9 5 \%}$ confidence bounds. Dots indicate observed values for catch, landings and discards. The vertical line in each plot delineates the last year of the historical assessment (2009): estimates to the right of these vertical lines are TSA-based forecasts.


Figure 3.3.5. Haddock in Division VIa. Standardised landings prediction errors from the final TSA run.


Figure 3.3.6. Haddock in Division VIa. Standardised discards prediction errors from the final TSA run.


Figure 3.3.7. Haddock in Division VIa. Standardised ScoGFS Q1 prediction errors from the final TSA run.


Figure 3.3.8. Haddock in Division VIa. Standardised ScoGFS Q4 prediction errors from the final TSA run.


Figure 3.3.9. Haddock in Division VIa. Stock-recruit plot from the final TSA run, points labelled as year classes. Predicted recruitments are circled: for the 2008 year-class recruiting in 2009 (using ScoGFS Q1 data); and the 2009 year-class recruiting in 2010 (based on the underlying Ricker model).


Figure 3.3.10. Haddock in Division VIa. Fitted (lines) and observed (dots) discard proportions-atage from the final TSA run.


Figure 3.3.11. Haddock in Division VIa. Estimates of Mean $F_{2-6}$, SSB and recruitment from retrospective TSA runs.


Figure 3.3.12. Haddock in Division VIa. Time-series of recruitment-at-age 1 from the final TSA assessment, along with the long-term (1978-2009) geometric mean and the age-1 indices from the Q1 and Q4 ScoGFS survey-series.


Figure 3.3.13. Haddock in Division VIa. Time-series of estimated fishing mortality-at-age, along with the mean over ages 2-6.


Figure 3.3.14. Haddock in Division VIa. Candidates for fishing mortality-at-age in short-term forecasts. Lines labelled 2005, 2006, 2007, 2008, 2009 indicate the TSA estimates for those years. Points marked 2009 TSA and 2010 TSA show the TSA-generated forecast values from the final assessment.

## Year class mean weights



Figure 3.3.15. Haddock in Division VIa. Mean weights-at-age (kg) in total catch (or stock), tracked by year class with a linear model fit. Predicted weights in 2010 based on linear model fits indicated with the dotted lines.

Figure Haddock,,,VIa,, Sensitivity analysis of short term forecast.


Figure 3.3.16. Haddock in Division VIa. Sensitivity analysis of short-term forecast.

Figure Haddock,VIa. Probability profiles for short term forecast.


Figure 3.3.17. Haddock in Division VIa. Probability profiles for short-term forecast.

Figure Haddock,VIa. Short term forecast


Figure 3.3.18. Haddock in Division VIa. Summary of short-term forecast.


Figure 3.3.19. Model fits from FMSY exploration based on the 2010 assessment.


Figure 3.3.20. Estimates of FMSY and other reference points based on the Smooth Hockey stick stock-recruitment relationship.
had-scow Ricker


Figure 3.3.21. Estimates of FMSY and other reference points based on the Ricker stockrecruitment function.

## had-scow Beverton-Holt



Figure 3.3.22. Estimates of FMSY and other reference points based on the Beverton and Holt stock-recruit function.


Figure 3.3.23. Reference points, yield-(removals) per-recruit and SSB per recruit analyses.

### 3.4. Whiting in Subarea Vla

Type of assessment in 2010
As agreed at this year's meeting of ACOM, assessment is being updated for whiting in Division VIa this year, following two years (2008 and 2009) when no advice was provided. Earlier, ACFM review groups (RGNSDS) highlighted the various data problems associated with this stock; including noisy survey data and discard data which need to be reworked. Their conclusion in 2006 was that:

Until revised Scottish discards are available and Irish discards included, a formal analytic assessment is not possible for this stock.

The assessment presented by the WG this year is therefore based only on survey data which is the same approach as that adopted in the 2007 assessment.

## ICES advice applicable to 2009 and 2010

In 2006, the ICES Advice for 2007 in terms of single stock exploitation boundaries was as follows:

Exploitation boundaries in relation to precautionary limits
"Given that SSB is estimated at the lowest observed level and total mortality at the highest level over the time period, catches in 2007 should be reduced to the lowest possible level."

The Advice given since then has been the same (see Table with the ICES Advice during 2001-2010 below).

### 3.4.1 General

## Stock description

General information is now located in the Stock Annex.

## Management applicable to 2009 and 2010

The TAC for whiting is set for ICES Subareas VI, XII and XIV and EU and international waters of ICES Subdivision Vb, and for 2010 was as shown below:

| Species: | Whiting <br> Meriangius merlangus | Zone:VL; EU and international waters of Vb; international <br> waters of XII and XIV <br> (WHG/561 214) |
| :--- | :--- | :--- | :--- |
| Germany | 3 |  |
| France | 53 |  |
| Ireland | 129 |  |
| United Kingdom | 246 | Analytical TAC |
| EU | 431 |  |
| TAC | 431 |  |

The following table summarises ICES advice and actual management applicable for whiting in Division VIa during 2001-2010:

| Year | Single species <br> EXPLOITATION (TONNES) | BASIS FOR SINGLE SPECIES | TAC For VB, VI, XII, XIV (tonnes) | \%CHANGE IN F <br> ASSOCIATED WITH TAC ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: |
| 2001 | <4200 | Reduce F below $\mathrm{F}_{\mathrm{p}} \mathrm{a}$ | 4000 | -40\% |
| 2002 | <2000 | SSB $>\mathrm{B}_{\mathrm{pa}}$ in short term | 3500 | -40\% |
| 2003 | - | $\mathrm{SSB}>\mathrm{B}_{\mathrm{pa}}$ in short term | 2000 | -60\% |
| 2004 | - | SSB $>\mathrm{B}_{\mathrm{pa}}$ in 2005 | 1600 | (no assessment) |
| 2005 | - | - | 1600 | (assessment in relative trends only) |
| 2006 | - | - | 1360 | (assessment in relative trends only) |
| 2007 | 0 | Reduce catches to lowest possible level | 1020 | (assessment in relative trends only) |
| 2008 | 0 | Reduce catches to lowest possible level | 765 | (no assessment) |
| 2009 | 0 | Reduce catches to lowest possible level | 574 | (no assessment) |
| 2010 | 0 | Reduce catches to lowest possible level | 431 | (assessment in relative trends only) |

${ }^{1}$ Based on F-multipliers from forecast tables.
The minimum landing size for whiting in Division VIa is 27 cm .

## Fishery in 2009

A description of the fisheries on the west of Scotland is given in Section 3.1.
Tables and figures of total effort to 2006 by the fleets operating in Division VIa can be found in Section 16 of the Report of WGNSDS 2007.

Anecdotal information from the fishing industry suggests that the number of vessels targeting whiting continues to be very low. However, the recent low TACs combined with increased interest in bigger whiting (driven by good prices) has resulted in an increasing uptake of the whiting quota. The quota for UK vessels in 2009 was slightly exceeded (by $9 \%$, compared to $84 \%$ and $49 \%$ of the quota taken up in 2008 in 2007, respectively, with post regulation quota swaps not being taken into account). Total landings in 2009 were 488 t , up slightly from 2008 (Table 3.4.1). These are above the lowest recorded landings of 2005, but continue to be far below the long-term average.

The total estimated international catch of ages $1-7+$ in 2009 was 905 t of which approximately 417 t were discards (Table 3.4. 2). An additional 417 t of $0-\mathrm{gp}$ fish were also estimated to be discarded. Although both the catch and discards in 2009 were higher than those in 2007 and 2008, they are still the third lowest in the respective time-series.

Mandatory introduction of larger square mesh panels for the Nephrops fleet in 2008 may be partially responsible for the relatively low catch and discards of whiting in

Division VIa. Despite the increase in discards in 2009, discarding is expected to remain low or to decline in subsequent years following the mandatory increase in mesh size to 120 mm for vessels fishing in the mixed demersal fishery in 2009.

### 3.4.2 Data

## Landings

Total landings, as officially reported to ICES in 1965-2009, are shown in Figure 3.4. 1. There have been concerns that the quality of landings data is deteriorating, giving a possible reason for the different stock dynamics implied by the commercial fleet and the annual survey (ScoGFS) in recent years (see Section 5.1.6.1.3 in the 2005 WG Report). The introduction of UK and Irish legislation requiring registration of all fish buyers and sellers may mean that the reported landings from 2006 onwards are more representative of actual landings.

Details on nations which supply data and sampling levels are given in Table 2.1. Age distributions were estimated from market samples. Annual numbers-at-age in the landings are given in Table 3.4.3. Annual mean weights-at-age in the landings are given in Table 3.4.6 and shown in Figure 3.4.2.

## Discards

Annual numbers-at-age in the discards are given in Table 3.4.4. Annual mean weights-at-age in the discards are given in Table 3.4.7 and shown in Figure 3.4.2.

This year, WG estimates of discards are based on data collected in the Irish and Scottish discard programme (raised by weighted average to the level of the total international discards). Discard age compositions from Scottish and Irish samples have been applied to unsampled fleets. Work is underway to revise the Scottish discard estimates with an aim to reduce bias and increase precision. Such revisions are particularly important for the estimation of total catch for this stock which has very high discards across a wide age range. A working document set out the methodology of this work at the 2004 meeting of WGNSDS (Fryer and Millar, 2004).

## Biological

Annual numbers-at-age in the total catch are given in Table 3.4.5. Annual mean weights-at-age in the total catch are given in Table 3.4.8. As in previous meetings, the catch mean weights-at-age were also used as stock mean weights-at-age (see Stock Annex).
Values for natural mortality ( 0.2 for all ages and years) and the proportion of fish ma-ture-at-age (knife-edged at age 2 for all years) are unchanged from the last assessment. Also as in the 2007 assessment, the proportion mature before spawning and the proportion fished before spawning are both set to be zero.

## Surveys

Four research survey indices for whiting in VIa were also available:

- Scottish west coast groundfish survey (ScoGFSQ1): ages 1-7, years 19852010.
- Irish west coast groundfish survey (IreGFS): ages 0-5, year 1993-2002.
- Scottish fourth-quarter west coast groundfish survey (ScoGFSQ4): ages 08, years 1996-2009.
- Irish groundfish survey (IRGFS): ages 0-6; years 2003-2009.

For the Scottish surveys, a new vessel and gear were used from 1999. The catch rates as presented are corrected for the change in vessel and gear. The basis for the correction is comparative trawl haul data (Zuur et al., 2001). The Irish quarter four survey was discontinued in 2003 and has been replaced by a new survey. The replacement survey (IRGFS) has been running for seven years. The Scottish quarter four survey was presented for the first time to WGNSDS 2005.

The survey-series are described in the Report of the 2009 IBTSWG and also in the Stock Annex. For all survey series, the oldest age given represents a true age, rather than a plus group. The survey indices are shown in Table 3.4 .9 with data used in the final assessment highlighted in bold. The sum over ages $1-7$ of the Scottish fourthquarter west coast groundfish survey indices amounted to $64 \%$ and the sum of the Scottish first-quarter west coast groundfish survey indices to $60 \%$ of the average in the respective time-series. Both sums were a two-fold increase over the previous year. The spatial distribution of cpue from the two Scottish surveys in 2009 and 2010 have been provided in the Stock Annex.

## Commercial cpue

Four commercial catch-effort dataseries were available to the WG including:

- Scottish light trawlers (ScoLTR): ages 1-7, years 1965-2005;
- Scottish seiners (ScoSEI): ages 1-6, years 1965-2005;
- Scottish Nephrops trawlers (ScoNTR): ages 1-6, years 1965-2005;
- Irish Otter Trawlers (IreOTB); ages 1-7, years 1995-2005.

Given the problems with non-mandatory effort reporting in the UK (described further in the report of WGNSSK for 2000, ICES CM 2001/ACFM:07), these cpue series have not been used for a number of years and are not presented in the Report. They are retained in the Stock Annex.

### 3.4.3 Historical stock development

The assessment is based only on survey data and is conducted using SURBA.

## Data screening and exploratory runs

Software used: SURBA 3.0
Model Options chosen: one or two tuning series used in one run
Input data types and characteristics:

- ScoGFSQ1: lambda=1, equal catchabilities at age, ages 1-6, all available years, mean $Z$ range 2-4
- ScoGFSQ4: lambda=1, equal catchabilities at age, ages 1-6, all available years, mean Z range 2-4

Software used: FLXSA 2.0
Input data types and characteristics:

- Catch data, ages 1-7+, years 1965-2009,
- ScoGFSQ1: lambda=1, equal catchabilities at age, ages 1-6, years 19952009.

Owing to uncertainties in catch-at-age data the WG only used commercial catch data to provide stock weights-at-age for this year's assessment.

Of the four survey-series available, only the two Scottish surveys were considered further. The new Irish survey (IRGFS) is relatively short (7 years data) to give useful information on stock trends while the Irish west coast groundfish survey (IreGFS) has been discontinued. In addition, the sub-sampling protocol of the IreGFS was altered mid-way through the survey and therefore there are doubts about the consistency of this series. The Irish series were therefore not considered further.

A comparison of scaled (standardised to z-scores) survey indices (from ScoGFSQ1 \& ScoGFSQ4) at age show similar trends for most ages (up to age 5, Figure 3.4.3).

Log mean-standardised survey indices by year class and by year and scatter-plots of indices within year classes are shown in Figures 3.4.4, 3.4.5 and 3.4.6. The year-class plots for both surveys are quite noisy and the ability of these surveys to reliably track year-class strength is generally poor. In addition, some of the correlations for the older ages in the ScoGFSQ1 scatterplot are negative, while the equivalent plots of the ScoGFSQ4 survey show very scattered data points. Age 0 in ScoGFSQ4 appears to be a particularly poor measure of year-class strength (little evidence of positive correlation) and is therefore excluded in further analysis of this survey. There are no marked year effects. The log catch curves for these surveys along with those for the catch are shown in Figure 3.4.7. The curves for both ScoGFSQ1 and ScoGFSQ4 are relatively linear and not very noisy, and show a fairly steep and consistent drop in abundance.
The trawl survey data (ScoGFSQ1 and ScoGFSQ4) for West of Scotland whiting were extensively analysed at WGNSDS 2005-2007 using both SURBA 2.2 and SURBA 3.0 to look at consistency of output using a variety of age ranges, smoothing parameter values, relative catchabilities and weighting factors. Initial single fleet SURBA runs this year therefore used the model settings that were chosen in 2007 with the extension of the age range for ScoGFS4 to 1-6 (as compared to ages 1-5 in the 2007 runs). This year only SURBA (version 3.0) was used to carry out the survey-based analysis; FLSURBA could not be run due to incompatibility of its available versions with the recent $R$ versions (in 2007, both SURBA and FLSURBA were run).

The summary output of mean $\mathrm{Z}(2-4)$, recruitment and biomass from the SURBA run for ScoGFSQ1 is shown in Figure 3.4.8 with the residuals illustrated in Figure 3.4.9. Model residuals are large for some age classes in some years, but with the exception of age 1 , do not show any particular trends or non-randomness. Little systematic retrospective bias is apparent in the stock trends although the estimates for recruitment show some variability (Figure 3.4.10). The mean $\mathrm{Z}(2-4)$ estimates from this run show large fluctuations over the examined period. Choosing larger values for the smoothing parameter (lambda) smoothed out the fluctuations in mean Z , but the runs showed much worse retrospective patterns (not shown).

The WG had some difficulty in applying the SURBA model to the ScoGFSQ4 survey in the 2007 runs. These problems were also present this year. The summary output for a run with the settings given above is shown in Figure 3.4.11 and the residuals in Figure 3.4.3.12. Some trends are similar to those obtained with the ScoGFSQ1 data. For total mortality, the trends are similar during 1996-2006 and after 2006, the trends are different. Model residuals are noisy, but show no particular trends or nonrandomness. No retrospective plots could be produced as some values were extremely high. The ScoGFSQ4 survey is a relatively short time series (in comparison to ScoGSQ1), without particularly good internal consistency or strong year-class signals and this may be the reason for the poor retrospective performance.

Since the two surveys appear to be implying consistent stock trends over some periods, a multifleet SURBA was also explored. The output was similar to that obtained with ScoGFSQ1 survey (not shown), but the residuals were noisy and over the year range when data are available from both surveys (1996-2009), there were some obvious problems with the model fit: virtually all residuals for the ScoGFSQ1 survey were negative while those for the ScoGFSQ4 survey were positive (not shown). The multifleet SURBA run was therefore not considered further.

In addition to SURBA runs, XSA was carried out with the ScoGFSQ1 survey. Despite the lack of independent discard estimates for the pre-1978 period, the whole catch data series (1965-2009) was used in the XSA run. The best performance was observed with the tuning series trimmed to 1995-2009. The output from XSA was compared with the SURBA run (for ScoGFSQ1) outputs, both being mean-standardised over the period 1995-2009 (Figure 3.4.13). There are substantial differences between the two outputs in the early period (from the mid-1980s to the mid-1990s), and there is more agreement between them from the mid-1990s onwards. Both models indicate a decline in mortality to low levels from 2004 to around the lowest in the time-series.

## Final update assessment

The SURBA run using ScoGFSQ1 data for ages $1-6$ is presented as the final assessment run given that it shows less retrospective problems than the ScoGFSQ4 survey. The SURBA model settings for the final run are given below:

Software used: SURBA 3.0
Model Options chosen: one tuning-series used
Input data types and characteristics:

- ScoGFSQ1: lambda=1, equal catchabilities at age, ages 1-6, all available years, mean Z range 2-4

These settings are the same as in the 2007 assessment. The output file from this run is given in Table 3.4.10. Trends in Z, recruitment and SSB from this run are shown in Figure 3.4.8. The residuals are shown in Figure 3.4 .9 and the retrospective in Figure 3.4.10. The level of SSB estimated in 2010 remains low and is comparable with that in 2009. Recruitment is estimated to have been very low in recent years, but the estimate for 2010 shows a considerable increase (by a factor of five) compared to 2009. Mean Z shows a decline from 2004 with relatively stable (though uncertain) levels in the recent years.

### 3.4.4 Short- term projections

No short-term predictions were made by this WG.

### 3.4.5 Medium- term projections

Stochastic medium-term predictions were not made at this WG because the assessment is considered only to be indicative of stock trends.

### 3.4.6 MSY explorations

No catch-based assessment was presented at the WG this year. The general lack of clear trends for the stock prevents using the final run output as the basis for advice.

### 3.4.7 Biological reference points

ICES considers that $B_{\lim }$ is $16000 t$ and $B_{p a}$ be set at 22000 t . ICES proposes that $\mathrm{F}_{\text {lim }}$ is 1.0 and $F_{p a}$ be set at 0.6.

The Working Group attempted a yield-per-recruit analysis with the output from the final SURBA run (Figure 3.4.14). $\mathrm{F}_{0.1}$ was estimated at around 0.3 and $\mathrm{F}_{\max }$ at around 0.5 , but it is unclear how stable these estimates are in the long term. The WG considers that yield-per-recruit F reference points are not applicable due to the uncertainty in historical stock trends.

### 3.4.8 Management plans

There are no specific management objectives or a management plan for this stock, but a plan is under development.

### 3.4.9 Uncertainties and bias in the assessment and forecast

The most significant problem with assessment of this stock is with commercial data. Incorrect reporting of landings (species and quantity) is known to occur and directly affects the perception of the stock. XSA is strongly influenced by estimated total catch data. Thus a survey-based assessment was used.

The survey data and commercial catch data contain different signals concerning the stock. The data since the mid-1990s are fairly consistent to conduct a catch-at-age analysis tuned with survey data. However, due to the discrepancy present in the earlier period, the Working Group considers that it is not possible to evaluate the current state of the stock with reference to precautionary reference points. A similar problem has been present in the North Sea whiting stock (as reported by WGNSSK 2010). Three potential sources of this discrepancy were identified for the North Sea stock, and they may apply to whiting in VIa as well: bias in catch estimates, changes in survey catchability or changes in natural mortality due to predation or regime shift (WGNSSK 2010).

Long-term information on the historical yield and catch composition indicates that the present stock size is low. The current assessment indicates (as the assessment carried out in 2007 did) that the stock is historically at a very low level. Total mortality has been declining over the past few years, but the most recent trends are unclear. The sum of the Scottish west coast groundfish survey indices (both in quarter one and quarter four) is also low, but shows an increase from 2008 onwards. The persistence of this trend should be verified in subsequent assessments.

### 3.4.10 Recommendation for next Benchmark

Catch-based assessment may potentially be a reliable basis for determining the status of the whiting stock in VIa. Currently, the main problem is the discrepancy between survey and catch data prior to 1995 . Unless this discrepancy can be resolved, truncating the catch data from 1995 may be an option, which proved satisfactory in exploratory XSA runs carried out at this working group. Given the new legislation on reporting landings, the quality of landings data is likely to continue to improve.
The potential for improvement in the quality of survey data needs to be investigated. The issue of changes in survey catchability needs to be addressed. The location of sampling stations may be reconsidered to better match the distribution of commercial landings.

### 3.4.11 Management considerations

Recruitment during the 1990s appears to have been high while more recently, it has been below average. There is an indication of a stronger 2009 year class following historically low recruitment of 2006 to 2008 year classes.

This year's assessment estimates SSB to remain at a low level, only marginally higher than the SSB observed in 2006-2008. Total mortality also remains low with uncertain developments. The perception of the state of this stock (as estimated from this assessment) appears not to have changed much, except for recruitment, from last year.

Whiting are caught in mixed fisheries with cod and haddock in VIa. Management of whiting will be strongly linked to that for cod for which there is an ongoing recovery plan (Council Regulation (EC) 1342/2008). There have also been several technical conservation measures introduced in the VIa gadoid fishery in recent years including the mandatory increases in mesh size to 120 mm .

Whiting are caught mainly as a bycatch species and there are no targeted fisheries for this stock, making direct management difficult. Whiting are caught and heavily discarded in small meshed fisheries for Nephrops. Any management measures which may result in a shift of vessels to these smaller mesh sizes will therefore result in a worse exploitation pattern and higher discards.

## 3.4b Whiting in Subarea VIb

Officially reported landings are given in Table 3.5.1.

## Table 3.4.1. Nominal landings ( t ) of WHITING in Division VIa, 1989-2009, as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1 | - | + | - | + | + | + | - | 1 | 1 | + | - | - | - | - | + | - | - | - | - | - |
| Denmark | 1 | + | 3 | 1 | 1 | + | + | + | + | - | - | - | - | - | + | + | - | - | - | - | - |
| Faroe <br> Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | + | + |
| France | 199 | 180 | 352 | 105 | 149 | 191 | 362 | 202 | 108 | 82 | 300 | 48 | 52 | 21 | 11 | 6 | 9 | 7 | 1 | 3 | 1 |
| Germany | + | + | + | 1 | 1 | + | - | + | - | - | + | - | - | - | - | - | - | + | 1 | - | - |
| Ireland | 1,315 | 977 | 1,200 | 1,377 | 1,192 | 1,213 | 1,448 | 1,182 | 977 | 952 | 1,121 | 793 | 764 | 577 | 568 | 356 | 172 | 196 | 56 | 69 | 125 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 2 |
| Spain | - | - | - | - | - | - | 1 | - | 1 | 2 | + | - | 2 | - | - | - | - | - | - | - | - |
| $\begin{aligned} & \text { UK (E, W \& } \\ & \text { NI) } \end{aligned}$ | 44 | 50 | 218 | 196 | 184 | 233 | 204 | 237 | 453 | 251 | 210 | 104 | 71 | 73 | 35 | 13 | 5 | 2 | 1 | - | - |
| UK (Scot.) | 6,109 | 4,819 | 5,135 | 4,330 | 5,224 | 4,149 | 4,263 | 5,021 | 4,638 | 3,369 | 3,046 | 2,258 | 1,654 | 1,064 | 751 | 444 | 103 | 178 | 424 | - | - |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 369 | 360 |
| Total landings | 7,669 | 6,026 | 6,908 | 6,010 | 6,751 | 5,786 | 6278 | 6642 | 6178 | 4657 | 4677 | 3203 | 2543 | 1735 | 1365 | 819 | 289 | 383 | 484 | 441 | 488 |

* Preliminary.

1989-2009 N. Ireland included with England and Wales.

Table 3.4.2. Whiting in Division VIa. Annual weight and numbers caught, years 1978-2009.

| Year | Weight (tonnes) |  |  | Numbers (thousands) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | Human consumption | Discards | Total | Human consumption | Discards |
| 1978 | 20452 | 14677 | 5775 | 93932 | 54369 | 39563 |
| 1979 | 20163 | 17081 | 3082 | 77794 | 61393 | 16401 |
| 1980 | 15108 | 12816 | 2292 | 57131 | 44562 | 12569 |
| 1981 | 16439 | 12203 | 4236 | 72113 | 46067 | 26046 |
| 1982 | 20064 | 13871 | 6193 | 87481 | 47883 | 39598 |
| 1983 | 21980 | 15970 | 6010 | 79114 | 49359 | 29755 |
| 1984 | 24118 | 16458 | 7660 | 125708 | 50218 | 75490 |
| 1985 | 23560 | 12893 | 10667 | 124683 | 43166 | 81517 |
| 1986 | 13413 | 8454 | 4959 | 64495 | 31273 | 33222 |
| 1987 | 18666 | 11544 | 7122 | 103485 | 41221 | 62264 |
| 1988 | 23136 | 11352 | 11784 | 141314 | 40681 | 100633 |
| 1989 | 11599 | 7531 | 4068 | 54633 | 26876 | 27757 |
| 1990 | 10036 | 5643 | 4393 | 42927 | 19201 | 23726 |
| 1991 | 12006 | 6660 | 5346 | 63112 | 25103 | 38009 |
| 1992 | 15396 | 6004 | 9392 | 86903 | 22266 | 64637 |
| 1993 | 15373 | 6872 | 8501 | 68351 | 23246 | 45105 |
| 1994 | 14771 | 5901 | 8870 | 87881 | 20060 | 67821 |
| 1995 | 13657 | 6076 | 7581 | 77932 | 18763 | 59169 |
| 1996 | 14058 | 7156 | 6902 | 71396 | 22329 | 49067 |
| 1997 | 11192 | 6285 | 4907 | 50459 | 19250 | 31209 |
| 1998 | 10476 | 4631 | 5845 | 56583 | 14387 | 42196 |
| 1999 | 7734 | 4613 | 3121 | 38260 | 15970 | 22290 |
| 2000 | 9715 | 3010 | 6705 | 78815 | 10118 | 68697 |
| 2001 | 4850 | 2438 | 2412 | 20802 | 8477 | 12325 |
| 2002 | 3829 | 1709 | 2120 | 25179 | 5765 | 19414 |
| 2003 | 2936 | 1356 | 1580 | 15403 | 4124 | 11279 |
| 2004 | 3437 | 811 | 2626 | 21749 | 2571 | 19178 |
| 2005 | 1239 | 341 | 898 | 6154 | 1051 | 5103 |
| 2006 | 1326 | 380 | 946 | 12988 | 1049 | 11939 |
| 2007 | 849 | 484 | 365 | 4879 | 1145 | 3734 |
| 2008 | 617 | 443 | 174 | 3085 | 1232 | 1853 |
| 2009 | 905 | 488 | 417 | 18038 | 1115 | 16923 |
| Min | 617 | 341 | 174 | 3085 | 1049 | 1853 |
| GM | 8207 | 4274 | 3484 | 44149 | 13944 | 26484 |
| AM | 11972 | 7067 | 4905 | 60399 | 24197 | 36203 |
| Max | 24118 | 17081 | 11784 | 141314 | 61393 | 100633 |

Table 3.4.3. Whiting in Division VIa. Landings-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 6938 | 6085 | 43530 | 4803 | 388 | 103 | 22 |
| 1966 | 1685 | 10544 | 2229 | 28185 | 1861 | 186 | 52 |
| 1967 | 5169 | 26023 | 10619 | 697 | 14574 | 789 | 143 |
| 1968 | 7265 | 16484 | 9239 | 3656 | 324 | 5036 | 368 |
| 1969 | 873 | 25174 | 8644 | 2566 | 1206 | 118 | 2333 |
| 1970 | 730 | 6423 | 28065 | 3241 | 670 | 214 | 550 |
| 1971 | 2387 | 8617 | 4122 | 34784 | 1338 | 240 | 223 |
| 1972 | 16777 | 12028 | 4013 | 1363 | 14796 | 793 | 148 |
| 1973 | 14078 | 36142 | 5592 | 1461 | 357 | 4292 | 310 |
| 1974 | 9083 | 51036 | 10049 | 1166 | 180 | 52 | 849 |
| 1975 | 14917 | 16778 | 36318 | 2819 | 281 | 57 | 245 |
| 1976 | 8500 | 46421 | 15757 | 17423 | 1508 | 66 | 57 |
| 1977 | 16120 | 13376 | 25144 | 3127 | 4719 | 292 | 24 |
| 1978 | 17670 | 18175 | 6682 | 9400 | 941 | 1433 | 68 |
| 1979 | 6334 | 34221 | 13282 | 3407 | 3488 | 276 | 384 |
| 1980 | 11650 | 11378 | 14860 | 4155 | 1244 | 1085 | 190 |
| 1981 | 3593 | 24395 | 11297 | 4611 | 1518 | 452 | 201 |
| 1982 | 2991 | 5783 | 29094 | 6821 | 2043 | 803 | 348 |
| 1983 | 3418 | 7094 | 8040 | 22757 | 6070 | 1439 | 540 |
| 1984 | 7209 | 12765 | 8221 | 4387 | 14825 | 1953 | 858 |
| 1985 | 4139 | 19520 | 8574 | 3351 | 1997 | 4764 | 822 |
| 1986 | 2674 | 14824 | 9770 | 2653 | 532 | 291 | 529 |
| 1987 | 6430 | 13935 | 13988 | 5442 | 837 | 330 | 259 |
| 1988 | 1842 | 20587 | 9638 | 6168 | 1949 | 290 | 207 |
| 1989 | 2529 | 5887 | 11889 | 4767 | 1266 | 468 | 71 |
| 1990 | 3203 | 8028 | 2393 | 4009 | 1326 | 204 | 37 |
| 1991 | 3294 | 8826 | 10046 | 1208 | 1391 | 286 | 51 |
| 1992 | 2695 | 9440 | 4473 | 4782 | 396 | 373 | 106 |
| 1993 | 1051 | 10179 | 6293 | 2673 | 2738 | 163 | 147 |
| 1994 | 909 | 4889 | 9158 | 3607 | 712 | 715 | 69 |
| 1995 | 215 | 4322 | 6516 | 5654 | 1397 | 376 | 282 |
| 1996 | 990 | 5410 | 7675 | 5052 | 2461 | 583 | 157 |
| 1997 | 877 | 3658 | 8514 | 4316 | 1441 | 338 | 106 |
| 1998 | 840 | 3504 | 4277 | 3698 | 1442 | 338 | 288 |
| 1999 | 1013 | 6131 | 4546 | 2040 | 1774 | 355 | 112 |
| 2000 | 484 | 2952 | 4211 | 1570 | 485 | 328 | 89 |
| 2001 | 461 | 3271 | 2630 | 1567 | 401 | 131 | 16 |
| 2002 | 62 | 1624 | 3018 | 799 | 227 | 23 | 13 |
| 2003 | 170 | 710 | 1111 | 1673 | 347 | 111 | 2 |
| 2004 | 54 | 724 | 543 | 521 | 622 | 78 | 29 |
| 2005 | 28 | 276 | 455 | 140 | 99 | 45 | 7 |
| 2006 | 82 | 139 | 369 | 260 | 61 | 113 | 24 |
| 2007 | 187 | 168 | 255 | 326 | 132 | 27 | 50 |
| 2008 | 6 | 265 | 394 | 336 | 152 | 55 | 24 |
| 2009 | 59 | 216 | 254 | 430 | 100 | 44 | 13 |

Table 3.4.4. Whiting in Division VIa. Discards-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 17205 | 4968 | 11437 | 531 | 14 | 2 | 0 |
| 1966 | 4322 | 8946 | 515 | 3317 | 79 | 3 | 0 |
| 1967 | 12237 | 20791 | 2674 | 84 | 629 | 12 | 1 |
| 1968 | 16394 | 12612 | 2137 | 377 | 13 | 82 | 3 |
| 1969 | 1983 | 20494 | 2093 | 292 | 51 | 2 | 26 |
| 1970 | 1776 | 6704 | 7494 | 382 | 33 | 4 | 0 |
| 1971 | 5505 | 6719 | 969 | 3906 | 57 | 4 | 1 |
| 1972 | 39192 | 8930 | 850 | 152 | 610 | 14 | 1 |
| 1973 | 30521 | 26995 | 1225 | 147 | 14 | 77 | 2 |
| 1974 | 23101 | 40590 | 2362 | 123 | 7 | 1 | 7 |
| 1975 | 37295 | 13541 | 8485 | 310 | 12 | 1 | 0 |
| 1976 | 24891 | 35812 | 3360 | 1940 | 63 | 1 | 0 |
| 1977 | 48148 | 8675 | 5432 | 301 | 212 | 5 | 0 |
| 1978 | 27942 | 10505 | 889 | 206 | 1 | 20 | 0 |
| 1979 | 3450 | 10722 | 1619 | 533 | 76 | 0 | 0 |
| 1980 | 2376 | 6172 | 3206 | 651 | 156 | 9 | 0 |
| 1981 | 1017 | 22014 | 2763 | 148 | 101 | 4 | 0 |
| 1982 | 17837 | 4577 | 15938 | 1189 | 55 | 1 | 0 |
| 1983 | 15069 | 8173 | 1964 | 4271 | 176 | 102 | 0 |
| 1984 | 68241 | 3951 | 1085 | 572 | 1577 | 59 | 4 |
| 1985 | 59783 | 17426 | 3134 | 663 | 61 | 446 | 3 |
| 1986 | 10459 | 20085 | 2491 | 117 | 6 | 2 | 61 |
| 1987 | 46876 | 13689 | 1518 | 180 | 1 | 0 | 0 |
| 1988 | 46421 | 51395 | 2472 | 292 | 54 | 0 | 0 |
| 1989 | 17778 | 3660 | 5796 | 401 | 111 | 11 | 0 |
| 1990 | 16406 | 5791 | 860 | 571 | 95 | 3 | 0 |
| 1991 | 30355 | 2874 | 4432 | 173 | 140 | 36 | 0 |
| 1992 | 46463 | 15041 | 2224 | 908 | 0 | 0 | 0 |
| 1993 | 14618 | 22281 | 5966 | 921 | 1317 | 0 | 2 |
| 1994 | 39697 | 18403 | 7775 | 1634 | 183 | 125 | 4 |
| 1995 | 28557 | 20921 | 8483 | 961 | 246 | 0 | 0 |
| 1996 | 28620 | 14617 | 4398 | 1395 | 18 | 1 | 18 |
| 1997 | 18182 | 9037 | 3431 | 466 | 93 | 0 | 0 |
| 1998 | 31183 | 7304 | 2418 | 991 | 184 | 51 | 64 |
| 1999 | 13623 | 7256 | 933 | 369 | 79 | 29 | 0 |
| 2000 | 63789 | 3556 | 1206 | 117 | 15 | 14 | 0 |
| 2001 | 5514 | 5861 | 738 | 208 | 4 | 0 | 0 |
| 2002 | 14166 | 3235 | 1749 | 130 | 124 | 8 | 1 |
| 2003 | 9331 | 1107 | 427 | 371 | 34 | 7 | 2 |
| 2004 | 14667 | 3557 | 536 | 305 | 107 | 4 | 2 |
| 2005 | 2923 | 1578 | 534 | 37 | 19 | 7 | 4 |
| 2006 | 9784 | 852 | 1000 | 256 | 36 | 11 | 2 |
| 2007 | 995 | 1077 | 308 | 64 | 4 | 3 | 0 |
| 2008 | 806 | 638 | 142 | 162 | 51 | 41 | 0 |
| 2009 | 6926 | 112 | 72 | 49 | 16 | 3 | 0 |

Table 3.4.5. Whiting in Division VIa. Total catch-at-age (thousands).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 24143 | 11054 | 54967 | 5334 | 402 | 105 | 22 |
| 1966 | 6007 | 19490 | 2744 | 31502 | 1940 | 189 | 53 |
| 1967 | 17406 | 46814 | 13293 | 781 | 15204 | 801 | 144 |
| 1968 | 23659 | 29096 | 11376 | 4034 | 337 | 5118 | 372 |
| 1969 | 2856 | 45668 | 10737 | 2858 | 1257 | 120 | 2358 |
| 1970 | 2506 | 13128 | 35559 | 3623 | 703 | 218 | 550 |
| 1971 | 7891 | 15336 | 5090 | 38690 | 1395 | 245 | 224 |
| 1972 | 55969 | 20958 | 4863 | 1514 | 15406 | 807 | 149 |
| 1973 | 44599 | 63137 | 6817 | 1608 | 371 | 4369 | 313 |
| 1974 | 32185 | 91625 | 12412 | 1289 | 188 | 53 | 856 |
| 1975 | 52213 | 30319 | 44804 | 3129 | 293 | 58 | 245 |
| 1976 | 33392 | 82233 | 19117 | 19363 | 1571 | 67 | 57 |
| 1977 | 64268 | 22051 | 30576 | 3428 | 4931 | 297 | 24 |
| 1978 | 45612 | 28680 | 7571 | 9606 | 942 | 1452 | 68 |
| 1979 | 9784 | 44943 | 14901 | 3940 | 3565 | 276 | 384 |
| 1980 | 14026 | 17551 | 18065 | 4806 | 1400 | 1093 | 190 |
| 1981 | 4610 | 46409 | 14060 | 4758 | 1618 | 456 | 201 |
| 1982 | 20829 | 10360 | 45032 | 8010 | 2098 | 804 | 348 |
| 1983 | 18487 | 15266 | 10004 | 27029 | 6246 | 1541 | 540 |
| 1984 | 75450 | 16716 | 9306 | 4959 | 16403 | 2011 | 863 |
| 1985 | 63922 | 36946 | 11708 | 4014 | 2058 | 5210 | 825 |
| 1986 | 13133 | 34909 | 12260 | 2770 | 539 | 293 | 591 |
| 1987 | 53305 | 27624 | 15506 | 5621 | 839 | 330 | 259 |
| 1988 | 48263 | 71982 | 12110 | 6460 | 2002 | 290 | 207 |
| 1989 | 20307 | 9547 | 17685 | 5168 | 1377 | 479 | 71 |
| 1990 | 19609 | 13819 | 3252 | 4580 | 1421 | 208 | 37 |
| 1991 | 33648 | 11700 | 14478 | 1381 | 1531 | 322 | 51 |
| 1992 | 49158 | 24481 | 6697 | 5691 | 396 | 373 | 106 |
| 1993 | 15669 | 32460 | 12259 | 3594 | 4055 | 163 | 149 |
| 1994 | 40606 | 23292 | 16933 | 5241 | 896 | 840 | 73 |
| 1995 | 28772 | 25243 | 14999 | 6615 | 1643 | 377 | 283 |
| 1996 | 29611 | 20027 | 12073 | 6447 | 2479 | 584 | 175 |
| 1997 | 19059 | 12695 | 11946 | 4782 | 1534 | 338 | 106 |
| 1998 | 32023 | 10808 | 6695 | 4689 | 1626 | 389 | 352 |
| 1999 | 14636 | 13387 | 5479 | 2408 | 1853 | 384 | 112 |
| 2000 | 64273 | 6508 | 5417 | 1687 | 500 | 343 | 89 |
| 2001 | 5975 | 9132 | 3368 | 1775 | 405 | 131 | 17 |
| 2002 | 14228 | 4859 | 4767 | 929 | 351 | 32 | 13 |
| 2003 | 9501 | 1817 | 1538 | 2044 | 381 | 119 | 4 |
| 2004 | 14721 | 4281 | 1079 | 825 | 730 | 82 | 31 |
| 2005 | 2951 | 1854 | 988 | 178 | 118 | 53 | 11 |
| 2006 | 9865 | 991 | 1369 | 516 | 97 | 124 | 26 |
| 2007 | 1182 | 1245 | 563 | 390 | 136 | 29 | 50 |
| 2008 | 812 | 903 | 536 | 498 | 203 | 96 | 24 |
| 2009 | 6985 | 328 | 325 | 478 | 116 | 47 | 13 |

Table 3.4.6. Whiting in Division VIa. Landings weights-at-age (kg).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.218 | 0.249 | 0.308 | 0.452 | 1.208 | 0.72 | 0.778 |
| 1966 | 0.238 | 0.243 | 0.325 | 0.374 | 0.61 | 0.72 | 0.828 |
| 1967 | 0.204 | 0.24 | 0.319 | 0.424 | 0.412 | 0.639 | 0.821 |
| 1968 | 0.206 | 0.263 | 0.366 | 0.444 | 0.554 | 0.538 | 0.735 |
| 1969 | 0.178 | 0.223 | 0.335 | 0.5 | 0.57 | 0.649 | 0.63 |
| 1970 | 0.205 | 0.203 | 0.274 | 0.382 | 0.519 | 0.619 | 0.683 |
| 1971 | 0.209 | 0.247 | 0.276 | 0.316 | 0.426 | 0.551 | 0.712 |
| 1972 | 0.211 | 0.258 | 0.345 | 0.368 | 0.426 | 0.494 | 0.638 |
| 1973 | 0.196 | 0.235 | 0.362 | 0.479 | 0.485 | 0.532 | 0.666 |
| 1974 | 0.193 | 0.215 | 0.317 | 0.444 | 0.591 | 0.641 | 0.584 |
| 1975 | 0.209 | 0.245 | 0.305 | 0.471 | 0.651 | 0.615 | 0.717 |
| 1976 | 0.201 | 0.242 | 0.309 | 0.361 | 0.497 | 0.687 | 0.856 |
| 1977 | 0.2 | 0.244 | 0.296 | 0.392 | 0.431 | 0.629 | 0.819 |
| 1978 | 0.199 | 0.235 | 0.286 | 0.389 | 0.516 | 0.549 | 0.612 |
| 1979 | 0.218 | 0.232 | 0.306 | 0.404 | 0.536 | 0.678 | 0.693 |
| 1980 | 0.172 | 0.242 | 0.33 | 0.42 | 0.492 | 0.595 | 0.817 |
| 1981 | 0.192 | 0.228 | 0.289 | 0.382 | 0.409 | 0.409 | 0.547 |
| 1982 | 0.184 | 0.22 | 0.276 | 0.352 | 0.505 | 0.513 | 0.526 |
| 1983 | 0.216 | 0.249 | 0.28 | 0.34 | 0.409 | 0.494 | 0.51 |
| 1984 | 0.216 | 0.259 | 0.313 | 0.371 | 0.412 | 0.458 | 0.458 |
| 1985 | 0.185 | 0.238 | 0.306 | 0.402 | 0.43 | 0.461 | 0.538 |
| 1986 | 0.174 | 0.236 | 0.294 | 0.365 | 0.468 | 0.482 | 0.499 |
| 1987 | 0.188 | 0.237 | 0.304 | 0.373 | 0.511 | 0.52 | 0.576 |
| 1988 | 0.176 | 0.215 | 0.301 | 0.4 | 0.483 | 0.567 | 0.6 |
| 1989 | 0.171 | 0.22 | 0.279 | 0.348 | 0.459 | 0.425 | 0.555 |
| 1990 | 0.225 | 0.251 | 0.324 | 0.359 | 0.417 | 0.582 | 0.543 |
| 1991 | 0.199 | 0.22 | 0.291 | 0.354 | 0.391 | 0.442 | 0.761 |
| 1992 | 0.193 | 0.23 | 0.288 | 0.349 | 0.388 | 0.397 | 0.51 |
| 1993 | 0.186 | 0.242 | 0.314 | 0.361 | 0.412 | 0.452 | 0.474 |
| 1994 | 0.161 | 0.217 | 0.29 | 0.371 | 0.451 | 0.482 | 0.483 |
| 1995 | 0.19 | 0.225 | 0.296 | 0.381 | 0.469 | 0.473 | 0.528 |
| 1996 | 0.195 | 0.245 | 0.288 | 0.365 | 0.483 | 0.526 | 0.569 |
| 1997 | 0.198 | 0.245 | 0.297 | 0.384 | 0.522 | 0.629 | 0.661 |
| 1998 | 0.215 | 0.236 | 0.301 | 0.364 | 0.438 | 0.5 | 0.646 |
| 1999 | 0.181 | 0.225 | 0.28 | 0.365 | 0.44 | 0.524 | 0.594 |
| 2000 | 0.205 | 0.241 | 0.298 | 0.336 | 0.419 | 0.488 | 0.617 |
| 2001 | 0.173 | 0.234 | 0.303 | 0.37 | 0.395 | 0.376 | 0.595 |
| 2002 | 0.213 | 0.257 | 0.304 | 0.363 | 0.464 | 0.65 | 0.707 |
| 2003 | 0.228 | 0.264 | 0.309 | 0.362 | 0.374 | 0.436 | 0.717 |
| 2004 | 0.193 | 0.251 | 0.295 | 0.345 | 0.382 | 0.403 | 0.342 |
| 2005 | 0.189 | 0.261 | 0.313 | 0.378 | 0.44 | 0.482 | 0.356 |
| 2006 | 0.221 | 0.292 | 0.319 | 0.394 | 0.455 | 0.528 | 0.567 |
| 2007 | 0.215 | 0.280 | 0.349 | 0.418 | 0.498 | 0.598 | 0.660 |
| 2008 | 0.274 | 0.245 | 0.322 | 0.384 | 0.514 | 0.530 | 0.653 |
| 2009 | 0.328 | 0.347 | 0.437 | 0.479 | 0.470 | 0.519 | 0.595 |

Table 3.4.7. Whiting in Division VIa. Discard weights-at-age (kg).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.122 | 0.177 | 0.213 | 0.249 | 0.287 | 0.303 | 0.287 |
| 1966 | 0.122 | 0.178 | 0.212 | 0.248 | 0.29 | 0.297 | 0.286 |
| 1967 | 0.122 | 0.178 | 0.213 | 0.248 | 0.29 | 0.295 | 0.289 |
| 1968 | 0.128 | 0.179 | 0.213 | 0.249 | 0.291 | 0.298 | 0.287 |
| 1969 | 0.121 | 0.178 | 0.214 | 0.249 | 0.29 | 0.295 | 0.285 |
| 1970 | 0.121 | 0.175 | 0.213 | 0.249 | 0.29 | 0.299 | 0.284 |
| 1971 | 0.12 | 0.177 | 0.211 | 0.248 | 0.29 | 0.299 | 0.284 |
| 1972 | 0.121 | 0.177 | 0.213 | 0.248 | 0.289 | 0.301 | 0.281 |
| 1973 | 0.123 | 0.176 | 0.215 | 0.252 | 0.288 | 0.301 | 0.285 |
| 1974 | 0.119 | 0.177 | 0.214 | 0.25 | 0.285 | 0.299 | 0.288 |
| 1975 | 0.119 | 0.176 | 0.213 | 0.25 | 0.286 | 0.301 | 0.278 |
| 1976 | 0.116 | 0.177 | 0.213 | 0.249 | 0.288 | 0.3 | 0.28 |
| 1977 | 0.118 | 0.177 | 0.214 | 0.249 | 0.289 | 0.299 | 0.282 |
| 1978 | 0.135 | 0.167 | 0.199 | 0.288 | 0.32 | 0.238 | 0 |
| 1979 | $0.173$ | 0.188 | 0.208 | 0.215 | 0.281 | 0 | 0 |
| 1980 | 0.14 | 0.179 | 0.208 | 0.22 | 0.271 | 0.386 | 0 |
| 1981 | $0.108$ | 0.16 | 0.195 | 0.298 | 0.286 | 0.295 | 0 |
| 1982 | 0.096 | 0.18 | 0.209 | 0.243 | 0.283 | 0.44 | 0 |
| 1983 | 0.141 | 0.186 | 0.228 | 0.237 | 0.267 | 0.267 | 0 |
| 1984 | 0.087 | 0.199 | 0.246 | 0.26 | 0.259 | 0.303 | 0.227 |
| 1985 | $0.102$ | 0.191 | 0.237 | 0.286 | 0.326 | 0.312 | 0.316 |
| 1986 | 0.092 | 0.17 | 0.196 | 0.245 | 0.258 | 0.33 | 0.263 |
| 1987 | 0.085 | 0.182 | 0.233 | 0.249 | 0.225 | 0 | 0 |
| 1988 | 0.076 | 0.143 | 0.203 | 0.227 | 0.262 | 0 | 0 |
| 1989 | 0.099 | 0.177 | 0.205 | 0.209 | 0.294 | 0.305 | 0 |
| 1990 | 0.124 | 0.171 | 0.214 | 0.219 | 0.237 | 0.264 | 0 |
| 1991 | 0.085 | 0.169 | 0.205 | 0.223 | 0.226 | 0.281 | 0 |
| 1992 | 0.109 | 0.173 | 0.219 | 0.227 | 0 | 0 | 0 |
| 1993 | 0.118 | 0.197 | 0.225 | 0.242 | 0.256 | 0 | 0.436 |
| 1994 | 0.087 | 0.157 | 0.22 | 0.283 | 0.297 | 0.253 | 0.299 |
| 1995 | 0.075 | 0.154 | 0.189 | 0.246 | 0.278 | 0.597 | 0.493 |
| 1996 | 0.095 | 0.18 | 0.203 | 0.229 | 0.302 | 0.421 | 0.26 |
| 1997 | 0.112 | 0.182 | 0.221 | 0.235 | 0.243 | 0.422 | 0.819 |
| 1998 | 0.098 | 0.179 | 0.225 | 0.254 | 0.282 | 0.264 | 0.245 |
| 1999 | 0.077 | 0.168 | 0.217 | 0.205 | 0.266 | 0.268 | 0 |
| 2000 | 0.075 | 0.164 | 0.203 | 0.233 | 0.282 | 0.25 | 0 |
| 2001 | 0.094 | 0.154 | 0.196 | 0.203 | 0.381 | 0 | 0 |
| 2002 | 0.073 | 0.162 | 0.212 | 0.245 | 0.24 | 0.295 | 0.276 |
| 2003 | 0.077 | 0.177 | 0.231 | 0.242 | 0.213 | 0.3 | 0.278 |
| 2004 | 0.086 | 0.186 | 0.236 | 0.246 | 0.304 | 0.349 | 0.314 |
| 2005 | 0.088 | 0.149 | 0.223 | 0.214 | 0.315 | 0.292 | 0.373 |
| 2006 | 0.046 | 0.197 | 0.235 | 0.295 | 0.322 | 0.518 | 0.362 |
| 2007 | 0.059 | 0.159 | 0.225 | 0.226 | 0.334 | 0.794 | 0.266 |
| 2008 | 0.075 | 0.211 | 0.286 | 0.301 | 0.397 | 0.222 | 0.304 |
| 2009 | 0.051 | 0.288 | 0.227 | 0.262 | 0.248 | 0.253 | 0 |

Table 3.4.8. Whiting in Division VIa. Total catch weights-at-age (kg).

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.15 | 0.217 | 0.288 | 0.432 | 1.177 | 0.712 | 0.776 |
| 1966 | 0.155 | 0.213 | 0.304 | 0.361 | 0.597 | 0.713 | 0.824 |
| 1967 | 0.146 | 0.212 | 0.298 | 0.405 | 0.407 | 0.634 | 0.817 |
| 1968 | 0.152 | 0.227 | 0.337 | 0.426 | 0.544 | 0.534 | 0.731 |
| 1969 | 0.138 | 0.203 | 0.311 | 0.474 | 0.559 | 0.643 | 0.626 |
| 1970 | 0.145 | 0.189 | 0.261 | 0.368 | 0.508 | 0.613 | 0.683 |
| 1971 | 0.147 | 0.216 | 0.264 | 0.309 | 0.42 | 0.547 | 0.71 |
| 1972 | 0.148 | 0.223 | 0.322 | 0.356 | 0.421 | 0.491 | 0.636 |
| 1973 | 0.146 | 0.21 | 0.336 | 0.458 | 0.477 | 0.528 | 0.663 |
| 1974 | 0.14 | 0.198 | 0.297 | 0.426 | 0.579 | 0.636 | 0.581 |
| 1975 | 0.145 | 0.214 | 0.288 | 0.449 | 0.636 | 0.61 | 0.717 |
| 1976 | 0.138 | 0.214 | 0.292 | 0.35 | 0.489 | 0.679 | 0.854 |
| 1977 | 0.139 | 0.218 | 0.281 | 0.379 | 0.425 | 0.624 | 0.816 |
| 1978 | 0.16 | 0.21 | 0.276 | 0.387 | 0.516 | 0.545 | 0.612 |
| 1979 | 0.202 | 0.222 | 0.295 | 0.378 | 0.531 | 0.678 | 0.693 |
| 1980 | 0.167 | 0.22 | 0.308 | 0.393 | 0.467 | 0.593 | 0.817 |
| 1981 | 0.173 | 0.196 | 0.271 | 0.379 | 0.401 | 0.408 | 0.547 |
| 1982 | 0.109 | 0.202 | 0.252 | 0.336 | 0.499 | 0.513 | 0.526 |
| 1983 | 0.155 | 0.215 | 0.27 | 0.324 | 0.405 | 0.479 | 0.51 |
| 1984 | 0.099 | 0.245 | 0.305 | 0.358 | 0.397 | 0.453 | 0.457 |
| 1985 | 0.107 | 0.216 | 0.288 | 0.383 | 0.427 | 0.448 | 0.537 |
| 1986 | 0.109 | 0.198 | 0.274 | 0.36 | 0.466 | 0.481 | 0.474 |
| 1987 | 0.097 | 0.21 | 0.297 | 0.369 | 0.51 | 0.52 | 0.576 |
| 1988 | 0.08 | 0.164 | 0.281 | 0.392 | 0.477 | 0.567 | 0.6 |
| 1989 | 0.108 | 0.204 | 0.255 | 0.337 | 0.446 | 0.422 | 0.555 |
| 1990 | 0.14 | 0.217 | 0.295 | 0.342 | 0.405 | 0.577 | 0.543 |
| 1991 | 0.096 | 0.207 | 0.265 | 0.338 | 0.376 | 0.424 | 0.761 |
| 1992 | 0.114 | 0.195 | 0.265 | 0.33 | 0.388 | 0.397 | 0.51 |
| 1993 | 0.123 | 0.211 | 0.271 | 0.331 | 0.361 | 0.452 | 0.474 |
| 1994 | 0.089 | 0.17 | 0.258 | 0.344 | 0.419 | 0.448 | 0.474 |
| 1995 | 0.076 | 0.166 | 0.235 | 0.361 | 0.44 | 0.473 | 0.528 |
| 1996 | 0.098 | 0.198 | 0.257 | 0.336 | 0.482 | 0.526 | 0.537 |
| 1997 | 0.116 | 0.2 | 0.275 | 0.369 | 0.505 | 0.629 | 0.661 |
| 1998 | 0.101 | 0.197 | 0.274 | 0.341 | 0.42 | 0.469 | 0.573 |
| 1999 | 0.084 | 0.194 | 0.269 | 0.34 | 0.433 | 0.504 | 0.593 |
| 2000 | 0.076 | 0.199 | 0.277 | 0.329 | 0.415 | 0.478 | 0.617 |
| 2001 | 0.1 | 0.183 | 0.28 | 0.35 | 0.395 | 0.376 | 0.589 |
| 2002 | 0.074 | 0.194 | 0.27 | 0.346 | 0.385 | 0.554 | 0.685 |
| 2003 | 0.08 | 0.211 | 0.287 | 0.34 | 0.36 | 0.427 | 0.526 |
| 2004 | 0.086 | 0.197 | 0.266 | 0.308 | 0.371 | 0.4 | 0.34 |
| 2005 | 0.089 | 0.166 | 0.264 | 0.344 | 0.42 | 0.455 | 0.362 |
| 2006 | 0.047 | 0.21 | 0.258 | 0.345 | 0.406 | 0.527 | 0.551 |
| 2007 | 0.084 | 0.175 | 0.281 | 0.387 | 0.494 | 0.616 | 0.659 |
| 2008 | 0.076 | 0.221 | 0.312 | 0.357 | 0.484 | 0.397 | 0.649 |
| 2009 | 0.053 | 0.327 | 0.391 | 0.457 | 0.440 | 0.500 | 0.572 |

Table 3.4.9. Whiting in Division VIa. Available survey tuning-series. Data used in final run are highlighted in bold. For ScoGFSQ1 and ScoGFSQ4, numbers are standardised to catch-rate per 10 hours. " + " indicates value less than 0.5 after standardising.

| SCOGFSQ1: Scottish Groundfish Sruvey - Effort in hours - Numbers- at- age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
| Year | (hours) | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1985 | 10 | 3140 | 1792 | 380 | 85 | 23 | 156 | 18 |
| 1986 | 10 | 1456 | 1526 | 403 | 68 | 10 | 9 | 10 |
| 1987 | 10 | 6938 | 1054 | 584 | 143 | 36 | 2 | 1 |
| 1988 | 10 | 567 | 3469 | 653 | 189 | 42 | 5 | 1 |
| 1989 | 10 | 910 | 505 | 586 | 237 | 48 | 3 | 0 |
| 1990 | 10 | 1818 | 572 | 122 | 216 | 61 | 4 | 1 |
| 1991 | 10 | 3203 | 277 | 298 | 22 | 39 | 9 | 1 |
| 1992 | 10 | 4777 | 1597 | 410 | 517 | 56 | 18 | 0 |
| 1993 | 10 | 5532 | 6829 | 644 | 91 | 30 | 11 | 2 |
| 1994 | 10 | 6614 | 2443 | 1487 | 174 | 56 | 15 | 6 |
| 1995 | 10 | 5598 | 2831 | 1160 | 370 | 70 | 17 | 32 |
| 1996 | 10 | 9384 | 2238 | 635 | 341 | 135 | 30 | 5 |
| 1997 | 10 | 5663 | 2444 | 1531 | 355 | 102 | 17 | 4 |
| 1998 | 10 | 9851 | 1352 | 294 | 195 | 50 | 14 | 1 |
| 1999 | 10 | 6125 | 4952 | 489 | 103 | 16 | 1 | 0.4 |
| 2000 | 10 | 12862 | 471 | 152 | 34 | 10 | 11 | 0 |
| 2001 | 10 | 4653 | 1954 | 242 | 41 | 8 | 1 | 1 |
| 2002 | 10 | 5542 | 1028 | 964 | 86 | 15 | 1 | 1 |
| 2003 | 10 | 6934 | 746 | 436 | 300 | 32 | 2 | 4 |
| 2004 | 10 | 5888 | 1566 | 189 | 131 | 44 | 9 | 1 |
| 2005 | 10 | 1308 | 723 | 183 | 35 | 8 | 11 | 2 |
| 2006 | 10 | 1441 | 466 | 282 | 77 | 0.3 | 3 | 0.6 |
| 2007 | 10 | 614 | 522 | 127 | 75 | 16 | 3 | 2 |
| 2008 | 10 | 593 | 127 | 77 | 26 | 8 | 3 | 0 |
| 2009 | 10 | 906 | 387 | 103 | 105 | 20 | 9 | 7 |
| 2010 | 10 | 3523 | 340 | 108 | 52 | 40 | 4 | 3 |


|  | IR- WCGFS : Irish West Coast GFS (VIa) - Effort in minutes - Numbers- at- age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |
| Year | (min) | 0 | 1 | 2 | 3 | 4 | 5 |
| 1993 | 2130 | 14403 | 32643 | 11419 | 1464 | 231 | 13 |
| 1994 | 1865 | 264 | 11969 | 4817 | 2812 | 78 | 57 |
| 1995 | 2026 | 34584 | 5609 | 6406 | 734 | 186 | 80 |
| 1996 | 2008 | 376 | 7457 | 3551 | 374 | 232 | 5 |
| 1997 | 1879 | 1550 | 13865 | 8207 | 1022 | 524 | 50 |
| 1998 | 1936 | 1829 | 4077 | 3361 | 663 | 121 | 5 |
| 1999 | 1914 | 3337 | 3059 | 1965 | 322 | 11 | 12 |
| 2000 | 1878 | 682 | 10102 | 2126 | 109 | 109 | 4 |
| 2001 | 965 | 1118 | 5201 | 2903 | 149 | 70 | 3 |
| 2002 | 796 | 594 | 8247 | 9348 | 820 | 280 | 0 |

Table 3.4.9 (continued).

|  | IRGFS: Irish groundfish survey - Effort in minutes - Numbers- at- age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
| Year | (min) | 0 | 1 | 2 | 3 | 4 | 5 | 6 |
| 2003 | 1127 | 1101 | 12886 | 2894 | 512 | 290 | 102 | 1 |
| 2004 | 1200 | 6924 | 3114 | 1312 | 104 | 35 | 16 | 1 |
| 2005 | 960 | 910 | 2228 | 1126 | 91 | 5 | 4 | 0 |
| 2006 | 1510 | 99 | 1055 | 921 | 214 | 27 | 3 | 0 |
| 2007 | 1173 | 138 | 1989 | 2380 | 722 | 169 | 251 | 122 |
| 2008 | 1135 | 24 | 4342 | 1328 | 573 | 243 | 123 | 36 |
| 2009 | 1378 | 16906 | 1430 | 989 | 325 | 68 | 21 | 41 |


|  | ScoGFSQ4 : Quarter four Scottish groundfish survey - Effort in hours - Numbers- at- age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |  |  |
| Year | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1996 | 10 | 5154 | 1908 | 1116 | 570 | 188 | 51 | 6 | 1 | 0 |
| 1997 | 10 | 8001 | 2869 | 951 | 323 | 160 | 46 | 12 | 1 | 0 |
| 1998 | 10 | 1852 | 2713 | 1124 | 149 | 100 | 20 | 1 | 0 | + |
| 1999 | 10 | 8203 | 2338 | 582 | 141 | 33 | 24 | 1 | 1 | 0 |
| 2000 | 10 | 4434 | 4055 | 789 | 160 | 9 | 7 | 1 | 0 | 0 |
| 2001 | 10 | 9615 | 1957 | 1420 | 155 | 40 | 12 | 2 | 0 | 0 |
| 2002 | 10 | 14658 | 1591 | 621 | 479 | 30 | 9 | 5 | 0 | 0 |
| 2003 | 10 | 9932 | 3446 | 567 | 338 | 83 | 27 | 4 | 0 | 0 |
| 2004 | 10 | 5923 | 1758 | 940 | 83 | 57 | 62 | 1 | 0 | 0 |
| 2005 | 10 | 2297 | 308 | 318 | 76 | 9 | 4 | 0.9 | 0.7 | 0 |
| 2006 | 10 | 415 | 296 | 140 | 101 | 35 | 8 | 3 | 0.5 | 0 |
| 2007 | 10 | 1894 | 434 | 326 | 99 | 83 | 48 | 0.6 | 0 | 0 |
| 2008 | 10 | 2297 | 208 | 78 | 110 | 28 | 24 | 4 | 0 | + |
| 2009 | 10 | 4833 | 236 | 178 | 50 | 58 | 12 | 6 | 6 | 0 |

Table 3.4.10. Whiting in Division VIa. Summary of SURBA indices of abundance-at-age, SSB and total mortality Z, based on data from ScoGFSQ1.

| Abundance- at- age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |
| Year | 1 | 2 | 3 | 4 | 5 | 6 |
| 1985 | 4.2996 | 1.4001 | 0.3581 | 0.0572 | 0.0332 | 0.168 |
| 1986 | 3.8295 | 1.41 | 0.3844 | 0.096 | 0.0114 | 0.007 |
| 1987 | 5.471 | 1.5089 | 0.4789 | 0.1279 | 0.0249 | 0.0031 |
| 1988 | 1.1208 | 2.1348 | 0.5067 | 0.1576 | 0.0327 | 0.0067 |
| 1989 | 1.7406 | 0.3821 | 0.613 | 0.1422 | 0.0331 | 0.0073 |
| 1990 | 1.2266 | 0.6442 | 0.1207 | 0.1895 | 0.0336 | 0.0082 |
| 1991 | 2.3483 | 0.523 | 0.2398 | 0.0441 | 0.0551 | 0.0102 |
| 1992 | 6.2786 | 1.5213 | 0.3161 | 0.1436 | 0.0235 | 0.03 |
| 1993 | 6.4522 | 3.1594 | 0.6861 | 0.1405 | 0.053 | 0.009 |
| 1994 | 5.1234 | 2.9258 | 1.2629 | 0.2696 | 0.0446 | 0.0176 |
| 1995 | 9.5743 | 2.2293 | 1.1149 | 0.4727 | 0.0807 | 0.0139 |
| 1996 | 6.8752 | 3.9321 | 0.7944 | 0.3897 | 0.1301 | 0.0233 |
| 1997 | 6.4495 | 2.5035 | 1.2188 | 0.2409 | 0.0901 | 0.0317 |
| 1998 | 8.2413 | 1.7058 | 0.5356 | 0.2534 | 0.035 | 0.014 |
| 1999 | 6.7661 | 1.8202 | 0.2961 | 0.09 | 0.0284 | 0.0042 |
| 2000 | 12.5175 | 1.4611 | 0.3078 | 0.0484 | 0.0097 | 0.0033 |
| 2001 | 4.0362 | 3.184 | 0.2988 | 0.0611 | 0.0067 | 0.0014 |
| 2002 | 1.8512 | 1.4346 | 0.9596 | 0.0881 | 0.0136 | 0.0016 |
| 2003 | 5.7885 | 0.725 | 0.4838 | 0.3172 | 0.0226 | 0.0037 |
| 2004 | 5.0915 | 1.8547 | 0.1937 | 0.1262 | 0.0609 | 0.0046 |
| 2005 | 1.5242 | 1.1384 | 0.3266 | 0.033 | 0.0144 | 0.0075 |
| 2006 | 1.3751 | 0.3631 | 0.2157 | 0.06 | 0.0041 | 0.0019 |
| 2007 | 0.5253 | 0.5093 | 0.1148 | 0.0668 | 0.0142 | 0.001 |
| 2008 | 0.5412 | 0.187 | 0.1538 | 0.0339 | 0.0149 | 0.0034 |
| 2009 | 0.7465 | 0.2983 | 0.0937 | 0.0761 | 0.0143 | 0.0065 |
| 2010 | 3.4327 | 0.3921 | 0.1414 | 0.0438 | 0.0299 | 0.0058 |

Table 3.4.10 (continued).

| Stock summary |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Rec |  | SSB | TSB | Mean Z(2-4) |  |
|  | Est | SElog | Est | Est | Est | SE |
| 1985 | 4.3 | 0.347 | 0.517 | 0.977 | 1.409 | 0.27 |
| 1986 | 3.83 | 0.312 | 0.428 | 0.845 | 1.177 | 0.206 |
| 1987 | 5.471 | 0.309 | 0.521 | 1.057 | 1.189 | 0.202 |
| 1988 | 1.121 | 0.315 | 0.574 | 0.663 | 1.36 | 0.199 |
| 1989 | 1.741 | 0.313 | 0.3 | 0.488 | 1.256 | 0.199 |
| 1990 | 1.227 | 0.311 | 0.258 | 0.43 | 1.077 | 0.2 |
| 1991 | 2.348 | 0.297 | 0.212 | 0.439 | 0.549 | 0.203 |
| 1992 | 6.279 | 0.303 | 0.449 | 1.158 | 0.868 | 0.201 |
| 1993 | 6.452 | 0.305 | 0.922 | 1.709 | 0.999 | 0.201 |
| 1994 | 5.123 | 0.306 | 0.943 | 1.399 | 1.051 | 0.2 |
| 1995 | 9.574 | 0.307 | 0.848 | 1.575 | 1.124 | 0.2 |
| 1996 | 6.875 | 0.31 | 1.184 | 1.858 | 1.276 | 0.201 |
| 1997 | 6.449 | 0.322 | 0.99 | 1.738 | 1.68 | 0.199 |
| 1998 | 8.241 | 0.331 | 0.59 | 1.423 | 1.908 | 0.197 |
| 1999 | 6.766 | 0.333 | 0.478 | 1.053 | 1.937 | 0.195 |
| 2000 | 12.517 | 0.327 | 0.398 | 1.349 | 1.73 | 0.197 |
| 2001 | 4.036 | 0.311 | 0.688 | 1.091 | 1.307 | 0.2 |
| 2002 | 1.851 | 0.307 | 0.574 | 0.711 | 1.184 | 0.201 |
| 2003 | 5.788 | 0.315 | 0.41 | 0.873 | 1.438 | 0.2 |
| 2004 | 5.091 | 0.333 | 0.48 | 0.918 | 1.893 | 0.198 |
| 2005 | 1.524 | 0.334 | 0.296 | 0.432 | 1.813 | 0.197 |
| 2006 | 1.375 | 0.319 | 0.155 | 0.22 | 1.255 | 0.2 |
| 2007 | 0.525 | 0.340 | 0.155 | 0.199 | 1.305 | 0.200 |
| 2008 | 0.541 | 0.358 | 0.110 | 0.151 | 0.753 | 0.202 |
| 2009 | 0.746 | 0.419 | 0.178 | 0.218 | 0.813 | 0.243 |
| 2010 | 3.433 | 0.562 | 0.175 | 0.419 | 0.957 | 0.114 |

Table 3.5.1. Nominal landings ( $\mathbf{t}$ ) of Whiting in Division VIb, 1989-2009, as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - |
| France | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | - | - | - | - | 32 | 10 | 4 | 23 | 3 | 1 | - | - | 10 |  | 2 | 3 | 3 | 104 | 16 | 23 | 23 |
| Spain | - | - | - | - | - | - | - | - | - | - | + | - | - | - | - | - | - | - | - | - | - |
| UK (E.\& W, NI) | 16 | 6 | 1 | 5 | 10 | 2 | 5 | 26 | 49 | 20 | + | + | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | 18 | 482 | 459 | 283 | 86 | 68 | 53 | 36 | 65 | 23 | 44 | 58 | 4 | 7 | 11 | 1 | 1 | 1 | 1 | $\ldots$ | $\ldots$ |
| UK (all) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 | 12 |
| Total | 34 | 488 | 460 | 288 | 128 | 80 | 62 | 85 | 117 | 44 | 44 | 58 | 14 | 7 | 13 | 4 | 4 | 105 | 17 | 31 | 35 |

* Preliminary.


Figure 3.4.1. Landings, discards and catch (in tonnes) of whiting in Division VIa, as officially reported to ICES.


Figure 3.4.2. Whiting in Division VIa. Mean weights-at-age in the landings (upper panel) and discards (lower panel).


Figure 3.4.3. Whiting in Division VIa. Comparison of scaled survey indices from ScoGFSQ1 and ScoGFSQ4.


Figure 3.4.4. Whiting in Division VIa. Log mean standardised survey index for each age by cohort (upper panel) and year (lower panel) in ScoGFSQ1.


Figure 3.4.5. Whiting in Division VIa. Log mean standardised survey index for each age by cohort (upper panel) and year (lower panel) in ScoGFSQ4.

## ScoGFSQ1


log index

## ScoGFSQ4


log index

Figure 3.4.6. Whiting in Division VIa. Comparative scatterplots at age for Scottish groundfish surveys, ScoGFSQ1 and ScoGFSQ4.


Figure 3.4.7. Whiting in Division VIa. Log catch curves from the catch (ages 1-7) and the two Scottish groundfish surveys, ScoGFSQ1 (ages 1-7) and ScoGFSQ4 (ages 0-7).

SURBA run with ScoGFSQ1 data


Figure 3.4.8. Whiting in Division VIa. Results of SURBA run using ScoGFSQ1 data. Mean total mortality estimates are given as absolute; biomass and recruitment are mean-standardised. Mean total mortality and recruitment are shown with $+/-$ standard errors.

## SURBA run with ScoGFSQ1 data - residuals



Figure 3.4.9. Whiting in Division VIa. Residuals by age from SURBA run using ScoGFSQ1 data.

## SURBA run with ScoGFSQ1 data



Figure 3.4.10. Whiting in Division VIa. Retrospective plots of SURBA run using ScoGFSQ1 data.

## SURBA run with ScoGFSQ4 data



Figure 3.4.11. Whiting in Division VIa. Results of SURBA run using ScoGFSQ4 data. Mean total mortality estimates are given as absolute; biomass and recruitment are mean-standardised. Mean total mortality and recruitment are shown with +/- standard errors.

## SURBA run with ScoGFSQ4 data - residuals



Figure 3.4.12. Whiting in Division VIa. Residuals by age from SURBA run using ScoGFSQ4 data.


Figure 3.4.13. Whiting in Division VIa. Comparison of trends based assessment final run outputs (SURBA) with VPA assessment (XSA) estimates. Fishing mortality, recruitment and SBB are mean-standardised over the period 1995-2009 (the length of the tuning-series used in XSA).


Figure 3.4.14. Whiting in Division VIa. Yield-per-recruit analysis with the output from the final SURBA run.

### 3.5 North Minch, FU1 1

Nephrops stocks have previously been identified by WGNEPH on the basis of population distribution, and defined as separate Functional Units. The Functional Units (FU) are defined by the groupings of ICES statistical rectangles given in Table 3.5.1 and illustrated in Figure 3.5.1. The Functional Unit is the level at which the WG collects fishery data (quantities landed and discarded, fishing effort, cpues and lpues, etc.) and length distributions, and at which it performs assessments.

There are three Functional Units in Division VIa, the level at which EU management of Nephrops currently takes place. Nominal landings as reported to ICES, along with WG estimates of landings are presented in Tables 3.5.2 and 3.5.3 respectively. Landings are also made from outside the Functional Units, from statistical rectangles where small pockets of suitable sediment exist, these are generally small amounts. There are no Functional Unit in Division VIb and only very small quantities of Nephrops are landed.

## Type of assessment in 2010

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG (WKNEPH, 2009) and described in Section 2.2.

### 3.5.1 Ecosystem aspects

The North Minch Functional Unit 11 at the northern end of the west coast of Scotland (Figure 3.5.1).

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the North Minch Functional Unit these substrates are distributed according to prevailing hydrographic and baythmetric conditions. The area is characterised by numerous islands of varying size and sea lochs occur along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of soft sediments. The North Minch exhibits the most patchy ground amongst west coast FUs. Very soft sediments are found in the southeast while coarser sandy muds prevail to the north and west. Figure 3.5.7 shows the distribution of sediment in the area.

Further information on ecosystem aspects can be found in the Stock Annex.

### 3.5.2 The fishery in 2009

The fishery in 2009 was generally similar to previous years with a fleet of mainly smaller trawlers working 1-4 day trips from the main ports of Lochinver, Ullapool, Stornaway and Gairloch. The largest part of the North Minch fleets continued to be based at Stornaway, made up of mostly smaller vessels, currently six single rigged trawlers and six muti-rigged trawlers, all but one are around 15 m length. The Barra fleet is more nomadic as the fishing grounds are more exposed which forces the fleet to find shelter on the east side of the North Minch. The Barra vessels are generally bigger than the Stornoway fleet, being all over 15 m in length. Although several vessels have been sold or left the fleet in recent years, the remainder have continued to fish the same pattern as always, most trawlers landing daily or every second day. In 2009 mesh size regulation went up from 80 to 90 mm . In the winter of 2009, high fuel prices and poor catches has resulted in boats not going out to fish. Under the west
coast emergency measures a square meshed panel of 120 mm was also required (Council Reg. (EU) 43/2009).

Little if any marketable fish bycatch was reported by the boats fishing in the North Minch, this was confirmed during Nephrops discard trips on board North Minch boats.

Further general information on the fishery can be found in the Stock Annex.

### 3.5.3 ICES advice in 2009

The ICES conclusions in 2009 in relation to State of the Stock were as follows:
"The stock is being exploited unsustainably. The UWTV survey indicates that the population has declined by around $40 \%$ over the past two years from a previous time-series high in 2006. Harvest ratios in this period were above the values associated with high long term yield and low risk of stock depletion."

The ICES advice for 2009 (Single-stock exploitation boundaries) was as follows:
"The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to the past two years (2006-2007). This corresponds to landings of no more than 4100 tonnes for the North Minch stock."

The ICES advice for 2010 (Single-stock exploitation boundaries) was as follows:
"ICES advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Rate for Nephrops fisheries should be less than $\mathrm{F}_{0.1}$. This corresponds to landings less than 972 t for the North Minch stock."

### 3.5.4 Management

Management is at the ICES subarea level as described at the beginning of Section 3.5.

### 3.5.5 Assessment

## Conclusions of the Review of the 2009 assessment

"RG agrees with the WG on the assessment and feels it follows the protocol described in the Stock Annex. The short-term projection gives various harvest rates and this should be used to assign the TAC. The idea of fishing at a level above $\mathrm{F}_{\text {max }}$ is unsettling and should be avoided especially for a stock that utilizes such a basic assessment."

## Approach in 2010

The assessment in 2010 is based on a combination of examining trends in fishery indicators and underwater TV using an extensive data-series for the North Minch.

The assessment of Nephrops and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG (WKNEPH 2009) and is described in Section 2.2. The provision of advice in 2010 develops the process defined by the Benchmark WG and described in Section 3.5 and attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES in 2010 (see Section 2.2). Intersessional work carried out by participants of the Benchmark and involving collaboration between WGNSSK and WGCSE is described in the working papers.

Previous TV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. Creel fishing is an important component of the North Minch fishery and landings from creel vessels have risen since the mid-1990s having been at a stable level since then. Given that creels operate across similar areas to those of the trawl fishery, this year's assessment is performed using combined length compositions from trawl and creels.

## Data available

An overview of the data provided and used by the WG is shown in Table 2.1.

## Commercial catch and effort data

Official catch statistics (landings) reported to ICES are shown in Table 3.5.2; these relate to the whole of VI of which the North Minch is a part. Landings by gear category for FU11 provided through national laboratories are presented in Table 3.5.5. Landings from this fishery are only reported from Scotland. A variety of gear types make landings of Nephrops. Total reported landings in 2009 were 3497 tonnes, consisting of 2858 tonnes landed by trawlers and 613 tonnes landed by creel vessels.

Given the concerns about the previously presented Scottish effort data (due to nonmandatory recording of hours fished in recent years) and following recommendations made by the RG, effort data in terms of days absent were presented to the WG. Reported effort by all Scottish Nephrops trawlers has shown a decreasing trend since 2002, (Figures 3.5.3 and 3.5.4).

The introduction of the "buyers and sellers" regulations in the UK in 2006 however, have led to increased reliability in the reported landings. Combined together, these observations imply that interpretation of lpue and cpue series is likely to be difficult and the increase in lpue after 2005 is probably reflecting the increase in reported landings rather than a change in stock abundance.

Males consistently make the largest contribution to the landings, although the sex ratio does seem to vary ( $79 \%$ males in 2009) (Figure 3.5.4). This is likely to be due to the varying seasonal pattern in the fishery and associated relative catchability (due to different burrow emergence behaviour) of male and female Nephrops. This occurs because males are available throughout the year and the fishery is also prosecuted in all quarters. Females on the other hand are mainly taken in the summer when they emerge after egg hatching.

Discarding of undersized and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 2000. Discarding rates in this FU average around $20 \%$ by number in the last five years It is likely that some Nephrops survive the discarding process, an estimate of $25 \%$ (Guéguen, J. and Chareau, A., 1975; Sangster et al., 1997; Wileman et al., 1999) survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard rate adjusted for survivorship estimated at the Benchmark Workshop was 19.9 \% (3 year average) and this value is used in the provision of landings options for 2011.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Quarterly landings and discards-at-length data were available from Scotland and these sampling levels
are shown in Table 3.5.4. Although assessments based on detailed catch analysis are not presently possible, examination of length compositions can provide a preliminary indication of exploitation effects.
Figure 3.5 .5 shows a series of annual length frequency distributions for the period 1979 to 2009. Catch (removals) length compositions are shown for each sex along with the mean size for both. In both sexes the mean sizes have been fairly stable over time. Examination of the tails of the distributions above 35 mm (the length beyond which the effects of recruitment pulses and discarding are considered to be negligible) shows no evidence of reductions in relative numbers of larger animals.

The observation of relatively stable length compositions is further confirmed in the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings (trawl only) shown in Figure 3.5.3 and Table 3.5.6. This parameter might be expected to reduce in size if overexploitation were taking place but there is no evidence of this. The mean size of smaller animals ( $<35 \mathrm{~mm}$ ) in the catch (and landings) is also quite stable through time up to 2009 where a slight drop can be observed for both males and females which may possibly indicate a good recruitment in this area. This result should however be interpreted with some caution since some inconsistent data was found in the length frequency distribution for the second quarter, when the catchability of females typically increases. To compensate for this, a fill-in from the others quarters was applied.

Mean weight in the landings is shown in Figure 3.5.6 and Table 3.5.9 and this also shows no systematic changes over the time-series.

## Natural mortality, maturity-at-age and other biological parameters

Biological parameter values are included in the Stock Annex.

## Research vessel data

Underwater TV surveys using a stratified random approach are available for this stock since 1994 (missing surveys in 1995 and 1997). Underwater television surveys of Nephrops burrow numbers and distributions, reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

The numbers of valid stations used in the final analysis in each year are shown in Table 3.5.8. On average, 38 stations have been considered valid each year, and then raised to the estimated area of the ground available for Nephrops- $1775 \mathrm{~km}^{2}$. In the 2009 TV survey there were fewer stations covering the usual strata (and therefore fewer used for the abundance calculation) because a number of exploratory stations were surveyed on the basis of newly available VMS data which indicated fishing activity in additional areas.

## Data analyses

## Exploratory analyses of survey data

A re-working of the UWTV survey abundances for Division VIa were presented to the Nephrops Benchmark Workshop (WKNEPH) in 2009 (ICES, 2009) and further details of the technical changes to the camera can be found in the Report of that workshop. The revised abundance estimates for FU11 from 1999 onwards were presented for the first time at WGCSE 2009 and are slightly higher than the previous values due to the field of view being smaller than previously calculated.

Table 3.5.7 shows the basic analysis for the three most recent TV surveys conducted in FU11. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach.

Figure 3.5.7 shows the distribution of stations in recent TV surveys (2004-2009), with the size of the symbols reflecting the Nephrops burrow density. Abundance is generally higher in the soft and intermediate sediments located to the south west and north east of the ground, however in previous years there has also been large abundances found on the coarser sediment type in the northeast of the. Table 3.5.8 and Figure 3.5.7 show the time-series estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates.

VMS plots (Figure 3.5.9) have shown fishing effort extends outside of the present survey area for FU11, which would imply an underestimate of stock biomass in this area. Further work needs to be done on the area estimate as the VMS data becomes more available.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential biases were highlighted including those due to edge effects, species burrow mis-identification and burrow occupancy. The cumulative bias correction factor estimated for FU11 was 1.33 meaning that the TV survey is likely to overestimate Nephrops abundance by $33 \%$.

## Final Assessment

The underwater TV survey is presented as the best available information on the North Minch Nephrops stock. The surveys provide a fishery independent estimate of Nephrops abundance. The details of the 2009 survey is shown in Table 3.5.7 and compared with the 2007 and 2008 outcome. At present it is not possible to extract any length or age structure information from the survey and it therefore only provides information on abundance over the area of the survey.

The 2009 TV survey data presented at this meeting shows that the abundance has increased slightly compared with 2008 (15\%) to the same level obtained for 2007 but the confidence limits overlap for the past three years.

The TV survey results reported here do not cover the sea loch areas adjacent to the main North Minch grounds and should therefore be considered underestimates of the overall biomass. The sea lochs support a significant but unknown percentage of the creel fishery. This issue is discussed further under quality of assessment.

### 3.5.6 Historic stock trends

The TV survey estimates of abundance for Nephrops in the North Minch suggest that historically the population increased until 2003 at which time it has fluctuated around the maximum value until 2006 when it declined for two years before a slight increase in 2009. The recently observed decrease has left stocks at a similar abundance to those seen in 2002 but not as low as previous to this. The bias adjusted abundance estimates from 1999-2009 (the period over which the survey estimates have been revised) are shown in Table 3.5.10. The stock is estimated to now be at 729 million individuals (bias adjusted values).

Table 3.5.10 also shows the estimated harvest ratios over this period. These range from $7-32 \%$. It is likely that prior to 2006 , the estimated harvest ratios may not be representative of actual harvest ratios due to under-reporting of landings).

### 3.5.7 MSY considerations

A number of potential $\mathrm{F}_{\mathrm{msy}}$ proxies are obtained from the per-recruit analysis for Nephrops and these are discussed further in Section 2.2 of this report. The analysis assumes the same input parameters (exploitation, discard ogive and biological parameters) as used at the Benchmark meeting in 2009. The complete range of the perrecruit $\mathrm{F}_{\text {msy }}$ proxies is given in the text table below and the process for choosing an appropriate $F_{\text {msy }}$ proxy is described in Section 2.2.
All $\mathrm{F}_{\mathrm{msy}}$ proxy harvest rate values are considered preliminary and may be modified following further data exploration and analysis.
For this FU, the absolute density observed on the UWTV survey is intermediate (based on the guideline categories suggested in Section 2.2) with an average of just over $0.55 \mathrm{~m}^{-2}$ suggesting the stock may have a medium productivity capability. Historical harvest ratios in this FU have been above that equivalent to fishing at $\mathrm{F}_{\max }$ and landings have been relatively stable in the last thirty years. $\mathrm{F}_{35 \% \mathrm{SpR}}$ (combined between sexes) is also estimated to be at $\mathrm{F}_{\text {max. }}$. For these reasons, the Working Group considered that F35\%SpR (combined between sexes) deliver high long-term yield with a low probability of recruitment overfishing and therefore is chosen as a proxy for $\mathrm{F}_{\text {msy }}$.

|  |  | Fbar(20-40 mm) |  |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fmult | M | F |  | M | F | T |
| $\mathrm{F}_{0.1}$ | M | 0.20 | 0.14 | 0.05 | 7.4 | 39.7 | 69.2 | 50.6 |
|  | F | 0.65 | 0.44 | 0.15 | 19.8 | 13.0 | 38.0 | 22.2 |
|  | T | 0.24 | 0.16 | 0.06 | 8.7 | 34.6 | 65.0 | 45.8 |
| $F_{\text {max }}$ | M | 0.36 | 0.24 | 0.08 | 12.2 | 24.3 | 54.4 | 35.4 |
|  | F | 1.49 | 1.01 | 0.34 | 37.2 | 4.7 | 18.2 | 9.6 |
|  | T | 0.52 | 0.35 | 0.12 | 16.6 | 16.7 | 44.2 | 26.8 |
| $\mathrm{F}_{35 \% \mathrm{SpR}}$ | M | 0.24 | 0.16 | 0.06 | 8.7 | 34.6 | 65.0 | 45.8 |
|  | F | 0.73 | 0.49 | 0.17 | 21.7 | 11.4 | 34.9 | 20.0 |
|  | T | 0.37 | 0.25 | 0.09 | 12.5 | 23.6 | 53.7 | 34.7 |

The Btrigger point for this FU (bias adjusted lowest observed UWTV abundance) is calculated as 330 million individuals.

### 3.5.8 Landings forecasts

A prediction of landings in 2011 based on principles established at the Benchmark Workshop WKNEPH (ICES 2009) and using the revised approach based on various proxies for FMsY (Dobby, 2009) outlined in the introductory Section 2.2 was made for the North Minch. The landings prediction for 2011 at the $\mathrm{F}_{\text {msy }}$ proxy harvest ratio is 1939 tonnes. Since current harvest rate is above the Fmsy proxy, the transition scheme towards the ICES MSY framework applies and would result in a landings estimate of 3118 tonnes for 2011.

The inputs to the landings forecast were as follows:
Mean weight in landings $(07-09)=24.2 \mathrm{~g}$
Discard rate (by number) $=12.1 \%$
Survey bias $=1.33$

|  | Harvest rate | Survey Index (ADJUSTED) | Implied fishery |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Retained number | Landings (tonnes) |
| $\mathrm{F}_{\text {ms }}$ | 12.5\% | 729 | 80 | 1939 |
| $\mathrm{F}_{\text {msy trasition }}$ | 20.1\% | 729 | 129 | 3118 |
| $\mathrm{F}_{0.1 \mathrm{M})}$ | 7.4\% | 729 | 47 | 1148 |
|  | 8.7\% | 729 | 56 | 1350 |
| $\mathrm{F}_{\text {max ( }}$ ) | 12.2\% | 729 | 78 | 1893 |
| F35\%spR(T) | 12.5\% | 729 | 80 | 1939 |
| $\mathrm{F}_{\text {max }}(\underline{1}$ | 16.6\% | 729 | 106 | 2575 |
| $\mathrm{F}_{2009}$ | 22.0\% | 729 | 141 | 3413 |

$\mathrm{F}_{0.1(\mathrm{M}, \mathrm{T}):}$ Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the male or combined sex YPR curve.
 bined SPR equal to $35 \%$ of the unfished level.
$\mathrm{F}_{\max (\mathrm{M}, \mathrm{T}) \text { : Harvest ratio equivalent to fishing at a rate which maximises the male or }}$ combined YPR.

A discussion of $\mathrm{F}_{\text {msy }}$ reference points for Nephrops is provided in Section 2.2.

### 3.5.9 Biological reference points

Precautionary approach biological reference points have not been determined for Nephrops stocks.

### 3.5.10 Unceratinties in the assessment and forecast

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the fishery adequately. In this assessment combined trawl and creel length compositions are used to account for the fact that the creel fishery accounts for over $17 \%$ of the landings, increasingly operates over similar areas to trawling, and exhibits a length composition composed of larger animals.

There were concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers was introduced and the reliability began to improve. Because of this the final assessment adopted is independent of official statistics. Harvest ratios since 2006 are also considered more reliable due to more accurate landings data reported under new legislation. Incorporation of creel length compositions has also improved estimates of harvest ratios.

Underwater TV surveys have been conducted for this stock since 1994, with a continual annual series available since 1998. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are quite small for this functional unit. There is a gap of 18 months between the survey and the start of the year for which the assessment is used to set management levels. It is assumed that the stock is in equilibrium during this period (i.e. recruitment and growth balance mortality) although this is rarely the case. The effect of this assumption on realised harvest rates has not been investigated.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three year average (2005-2007) of discard rate (adjusted to account for some sur-
vival of discarded animals) has been used in the calculation of catch options. The recent observed discard rate shows a large decline in discards in 2007 and 2008 coincident with a drop in survey abundance but a return to former levels in 2009.

The cumulative bias estimates for FU11 are largely based on expert opinion (See Annex). The precision of these bias corrections cannot yet be characterised.

The overall area of the ground is estimated by contoured sediment data. New VMS data linked to landings (through interrogation of the Scottish FIN system) suggests that not all areas are being considered in the current UWTV approach and as such, the absolute abundance estimate for this ground is likely to be an underestimate. Figure 3.5.9 illustrates differences between the British Geological Survey based sediment approach to estimating area and the activity of $>15 \mathrm{~m}$ trawlers; inclusion of smaller vessels would likely further modify this. Work is in progress to refine the area estimate.

### 3.5.11 Status of the stock

The perception of the state of the stock has not changed substantially since the assessment in 2008. The evidence from the TV survey suggests that the population is stable, but at a lower level than that evident from 2003-2006 and the $14 \%$ increase observed in 2009 is within the confidence limits for the past two years. The calculated harvest ratio in 2009 (dead removals/TV abundance) is above the values associated with high long-term yield and low risk depletion.

### 3.5.12 Management considerations

The WG, ACFM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level and management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort were compatible and in line with the scale of the resource.

Creel fishing takes place in this area but overall effort by this fleet in terms of creel numbers is not known and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the North Minch and STECF estimates that discards of whiting and haddock are high in VIa generally. It is important that efforts are made to ensure that unwanted bycatch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod under the Scottish Conservation Credits scheme and west coast emergency measures include the implementation of larger meshed square meshed panels ( 120 mm ) and real time closures to avoid cod.

The implementation of buyers and sellers legislation in the UK in 2006 is improving the reliability of fishery statistics but the transition period is accompanied in some cases by large changes in landings which produce significant changes in the lpue and cpue series that cannot be completely attributed to changes in stock. Until a sufficient time-series of reliable data has built up, use of fishery catch and effort data in the assessment process should be avoided.

### 3.5.13 References

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Table 3.5.1. Nephrops Functional Units and descriptions by statistical rectangle.

| Functional <br> UNIt | Stock | Division | ICES Rectanales |
| :---: | :--- | :---: | :--- |
| 11 | North Minch | Vla | $44-46$ E3-E4 |
| 12 | South Minch | VIa | $41-43$ E2-E4 |
| 13 | Clyde | Vla | $39-40$ E4-E5 |
| 14 | Irish Sea East | VIla | $35-38 E 6 ; 38 E 5$ |
| 15 | Irish Sea West | VIla | $36 E 3 ; 35-37$ E4-E5; 38E4 |

Table 3.5.2. Nominal catch (tonnes) of Nephrops in Division VIa and VIb, 1980-2009, as officially reported to ICES. There are no Functional Units in ICES Division VIb but occasional small landings are made.

Via Official Landings

|  | France | IRELAND | Spain | UK-(EnGL+WALES+N.IRL) | UK- Scotland | UK | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 5 | 1 | - | - | 7,422 | - | 7,428 |
| 1981 | 5 | 26 | - | - | 9,519 | - | 9,550 |
| 1982 | 1 | 1 | - | 1 | 9,000 | - | 9,003 |
| 1983 | 1 | 1 | - | 11 | 10,706 | - | 10,719 |
| 1984 | 3 | 6 | - | 12 | 11,778 | - | 11,799 |
| 1985 | 1 | 1 | 28 | 9 | 12,449 | - | 12,488 |
| 1986 | 8 | 20 | 5 | 13 | 11,283 | - | 11,329 |
| 1987 | 6 | 128 | 11 | 15 | 11,203 | - | 11,363 |
| 1988 | 1 | 11 | 7 | 62 | 12,649 | - | 12,730 |
| 1989 | - | 9 | 2 | 25 | 10,949 | - | 10,985 |
| 1990 | - | 10 | 4 | 35 | 10,042 | - | 10,091 |
| 1991 | - | 1 | - | 37 | 10,458 | - | 10.496 |
| 1992 | - | 10 | - | 56 | 10,783 | - | 10,849 |
| 1993 | - | 7 | - | 191 | 11,178 | - | 11,376 |
| 1994 | 3 | 6 | - | 290 | 11,047 | - | 11,346 |
| 1995 | 4 | 9 | 3 | 346 | 12,527 | - | 12,889 |
| 1996 | - | 8 | 1 | 176 | 10,929 | - | 11,114 |
| 1997 | - | 5 | 15 | 133 | 11,104 | - | 11,257 |
| 1998 | - | 25 | 18 | 202 | 10,949 | - | 11,194 |
| 1999 | - | 136 | 40 | 256 | 11,078 | - | 11,510 |
| 2000 | 1 | 130 | 69 | 137 | 10,667 | - | 11,004 |
| 2001 | 9 | 115 | 30 | 139 | 10,568 | - | 10,861 |
| 2002 | - | 117 | 18 | 152 | 10,225 | - | 10,512 |
| 2003 | - | 145 | 12 | 81 | 10,450 | - | 10,688 |
| 2004 | - | 150 | 6 | 267 | 9,941 | - | 10,364 |
| 2005 | - | 153 | 17 | 153 | 7,616 | - | 7,939 |
| 2006 | - | 133 | 1 | 255 | 13,432 | - | 13,821 |
| 2007 | - | 155 | - | 2,088 | 14,120 | - | 16,363 |
| 2008 | - | 56 | 1 | 419 | 14,795 | - | 15,271 |
| 2009* | - | 56 | - | - | - | 12,634 | 12,690 |

[^2]VIb Official Landings

|  | France | Germany | Ireland | Spain | UK-(EnGL+WALES+N.IRL) | UK-SCOTLAND | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | - | - | - | - | - | - | 0 |
| 1981 | - | - | - | - | - | - | 0 |
| 1982 | - | - | - | - | - | - | 0 |
| 1983 | - | - | - | - | - | - | 0 |
| 1984 | - | - | - | - | - | - | 0 |
| 1985 | - | - | - | - | - | - | 0 |
| 1986 | - | - | - | 8 | - | - | 8 |
| 1987 | - | - | - | 18 | 11 | - | 29 |
| 1988 | - | - | - | 27 | 4 | - | 31 |
| 1989 | - | - | - | 14 | - | - | 14 |
| 1990 | - | - | - | 10 | 1 | - | 11 |
| 1991 | - | - | - | 30 |  | - | 30 |
| 1992 | - | - | - | 2 | 4 | 1 | 7 |
| 1993 | - | - | - | 2 | 6 | 9 | 17 |
| 1994 | - | - | - | 5 | 16 | 5 | 26 |
| 1995 | 1 | - | - | 2 | 26 | 1 | 30 |
| 1996 | - | 6 | - | 5 | 65 | 5 | 81 |
| 1997 | - | - | 1 | 3 | 88 | 23 | 115 |
| 1998 | - | - | 1 | 6 | 46 | 7 | 60 |
| 1999 | - | - | - | 5 | 2 | 5 | 12 |
| 2000 | 2 | - | 8 | 3 | 4 | 4 | 21 |
| 2001 | 1 | - | 1 | 14 | 2 | 7 | 25 |
| 2002 | 1 | - | - | 7 | 3 | 7 | 18 |
| 2003 | - | - | 1 | 5 | 6 | 18 | 30 |
| 2004 | - | - | - | 2 | 7 | 13 | 22 |
| 2005 | 3 | - | 1 | 1 | 5 | 7 | 17 |
| 2006 | - | - | - | - | 1 | 3 | 4 |
| 2007 | - | - | - | 2 | - | - | 2 |
| 2008 | - | - | - | - | - | - | 0 |
| 2009* | - | - | - | - | - | - | 0 |

* figures are provisional.

Table 3.5.3. Nephrops, Total Nephrops landings (tonnes) by Functional Unit plus Other rectangles, 1981-2009.

| Year | FU11 | FU12 | FU13 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 2861 | 3651 | 2968 | 39 | 9519 |
| 1982 | 2799 | 3552 | 2623 | 27 | 9001 |
| 1983 | 3196 | 3412 | 4077 | 34 | 10719 |
| 1984 | 4144 | 4300 | 3310 | 36 | 11790 |
| 1985 | 4061 | 4008 | 4285 | 104 | 12458 |
| 1986 | 3382 | 3484 | 4341 | 89 | 11296 |
| 1987 | 4083 | 3891 | 3007 | 257 | 11238 |
| 1988 | 4035 | 4473 | 3665 | 529 | 12702 |
| 1989 | 3205 | 4745 | 2812 | 212 | 10974 |
| 1990 | 2544 | 4430 | 2912 | 182 | 10068 |
| 1991 | 2792 | 4442 | 3038 | 255 | 10527 |
| 1992 | 3560 | 4237 | 2805 | 248 | 10849 |
| 1993 | 3192 | 4455 | 3342 | 344 | 11332 |
| 1994 | 3616 | 4415 | 2629 | 441 | 11101 |
| 1995 | 3656 | 4680 | 3989 | 460 | 12785 |
| 1996 | 2871 | 3995 | 4060 | 239 | 11165 |
| 1997 | 3046 | 4345 | 3618 | 243 | 11252 |
| 1998 | 2441 | 3730 | 4843 | 157 | 11171 |
| 1999 | 3257 | 4051 | 3752 | 438 | 11498 |
| 2000 | 3246 | 3952 | 3419 | 421 | 11038 |
| 2001 | 3259 | 3992 | 3182 | 420 | 10853 |
| 2002 | 3440 | 3305 | 3383 | 397 | 10525 |
| 2003 | 3268 | 3879 | 3171 | 433 | 10751 |
| 2004 | 3135 | 3868 | 3025 | 403 | 10431 |
| 2005 | 2984 | 3841 | 3423 | 254 | 10502 |
| 2006 | 4160 | 4554 | 4778 | 241 | 13733 |
| 2007 | 3968 | 5451 | 6656 | 259 | 16334 |
| 2008 | 3799 | 5347 | 5921 | 162 | 15229 |
| 2009* | 3497 | 4267 | 4405 | 171 | 12340 |

* provisional.

Table 3.5.4. Nephrops. Sampling levels all FUs in VIa.

| IMS DATA ONLY | 2007 | 2008 | 2009* |
| :--- | :---: | :---: | :---: |
| No. Nephrops Samples | 126 | 119 | 144 |
| No. Nephropsmeasured | 119962 | 68309 | 81692 |
|  |  |  |  |
| DIscard DATA ONLY | 2007 | 2008 | $2009 *$ |
| No. Nephrops Samples | 22 |  | 18 |
| No. Marketable Nephropsmeasured | NA | 45251 | 32663 |
| No. Discards Measured | 14630 | 15975 | 17833 |

* 2009 is not directly comparable with previous years given that sampling levels shown are aggregated for all gears while sampling numbers for 2007 and 2008 include only Nephrops trawl and Creel fishing.

Table 3.5.5. Nephrops, North Minch (FU11), Nominal Landings of Nephrops, 1981-2009.

| Year | UK Scotland |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Nephropstrawl | Other trawl | Creel | Total |
| 1981 | 2320 | 170 | 371 | 2861 |
| 1982 | 2323 | 105 | 371 | 2799 |
| 1983 | 2784 | 95 | 317 | 3196 |
| 1984 | 3449 | 161 | 534 | 4144 |
| 1985 | 3236 | 117 | 708 | 4061 |
| 1986 | 2642 | 203 | 537 | 3382 |
| 1987 | 3458 | 143 | 482 | 4083 |
| 1988 | 3449 | 149 | 437 | 4035 |
| 1989 | 2603 | 112 | 490 | 3205 |
| 1990 | 1941 | 134 | 469 | 2544 |
| 1991 | 2228 | 125 | 439 | 2792 |
| 1992 | 2978 | 150 | 432 | 3560 |
| 1993 | 2699 | 85 | 408 | 3192 |
| 1994 | 2916 | 246 | 454 | 3616 |
| 1995 | 2940 | 184 | 532 | 3656 |
| 1996 | 2355 | 147 | 369 | 2871 |
| 1997 | 2553 | 102 | 391 | 3046 |
| 1998 | 2023 | 67 | 351 | 2441 |
| 1999 | 2791 | 56 | 410 | 3257 |
| 2000 | 2695 | 28 | 523 | 3246 |
| 2001 | 2651 | 41 | 567 | 3259 |
| 2002 | 2775 | 79 | 586 | 3440 |
| 2003 | 2607 | 44 | 617 | 3268 |
| 2004 | 2400 | 25 | 710 | 3135 |
| 2005 | 2267 | 18 | 699 | 2984 |
| 2006 | 3446 | 17 | 697 | 4160 |
| 2007 | 3362 | 16 | 590 | 3968 |
| 2008 | 3230 | 12 | 557 | 3799 |
| 2009 | 2858 | 26 | 613 | 3497 |

* provisional na = not available
** There are no landings by other countries from this FU

Table 3.5.6. Nephrops, North Minch (FU11): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 1981-2009.

| Year | Catches |  | LANDINGS |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | < 35 mm CL |  | > 35 mm CL |  |
|  | Males | Females | Males | Females | Males | Females |
| 1981 | 30.2 | 29.3 | 30.6 | 30.2 | 39.2 | 37.6 |
| 1982 | 29.8 | 28.6 | 30.1 | 29.0 | 39.8 | 37.4 |
| 1983 | 29.0 | 27.6 | 29.1 | 27.5 | 40.0 | 37.8 |
| 1984 | 28.5 | 28.0 | 28.5 | 28.1 | 39.2 | 37.4 |
| 1985 | 27.9 | 27.5 | 27.9 | 27.5 | 40.0 | 37.5 |
| 1986 | 29.5 | 28.4 | 29.7 | 28.6 | 39.1 | 37.6 |
| 1987 | 29.6 | 29.0 | 29.9 | 29.6 | 39.8 | 37.9 |
| 1988 | 29.9 | 29.5 | 30.3 | 30.1 | 38.9 | 38.0 |
| 1989 | 29.0 | 29.0 | 29.2 | 29.2 | 40.1 | 38.9 |
| 1990 | 29.3 | 28.6 | 29.8 | 28.9 | 39.1 | 38.1 |
| 1991 | 30.3 | 29.1 | 30.6 | 29.5 | 39.4 | 39.1 |
| 1992 | 29.3 | 28.0 | 29.7 | 28.3 | 39.6 | 38.3 |
| 1993 | 29.4 | 27.9 | 29.5 | 28.0 | 38.7 | 38.3 |
| 1994 | 28.1 | 27.0 | 29.4 | 28.3 | 39.5 | 38.8 |
| 1995 | 27.7 | 27.7 | 28.6 | 29.0 | 40.0 | 38.2 |
| 1996 | 29.5 | 29.4 | 30.2 | 30.2 | 40.0 | 38.7 |
| 1997 | 29.1 | 28.4 | 29.9 | 28.8 | 39.4 | 38.0 |
| 1998 | 29.8 | 28.8 | 30.6 | 29.3 | 39.6 | 38.4 |
| 1999 | 28.9 | 28.2 | 30.1 | 29.1 | 39.4 | 37.5 |
| 2000 | 29.9 | 28.6 | 30.4 | 29.0 | 39.4 | 37.8 |
| 2001 | 29.4 | 28.1 | 30.3 | 28.8 | 39.8 | 38.2 |
| 2002 | 29.2 | 28.4 | 30.4 | 29.5 | 39.7 | 38.3 |
| 2003 | 29.0 | 28.3 | 30.3 | 29.6 | 39.2 | 37.8 |
| 2004 | 29.6 | 28.9 | 30.4 | 29.5 | 40.3 | 38.8 |
| 2005 | 28.4 | 27.8 | 30.1 | 30.0 | 39.4 | 37.8 |
| 2006 | 29.0 | 27.4 | 30.5 | 28.9 | 39.1 | 38.2 |
| 2007 | 30.0 | 28.3 | 30.0 | 28.2 | 40.3 | 38.7 |
| 2008 | 29.6 | 28.3 | 30.1 | 28.8 | 40.0 | 38.5 |
| 2009 | 27.9 | 25.1 | 28.9 | 25.3 | 39.4 | 38.3 |

[^3]Table 3.5.7. Nephrops, North Minch (FU11): Results by stratum of the 2007-2009 TV surveys. Note that stratification was based on a series of arbitrary rectangles ( $\mathrm{U}, \mathrm{V}, \mathrm{W}, \mathrm{X}$ ).

|  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 TV survey |  |  |  |  |  |  |  |
| U | 656 | 14 | 0.53 | 0.11 | 348 | 3475 | 0.407 |
| V | 425 | 9 | 0.70 | 0.12 | 296 | 2328 | 0.273 |
| W | 563 | 9 | 0.25 | 0.07 | 142 | 2319 | 0.272 |
| X | 131 | 4 | 0.92 | 0.10 | 121 | 412 | 0.048 |
| Total | 1775 | 36 |  |  | 907* | 8534 | 1 |
| 2008 TV survey |  |  |  |  |  |  |  |
| U | 656 | 13 | 0.36 | 0.05 | 233 | 1511 | 0.255 |
| V | 425 | 10 | 0.59 | 0.05 | 250 | 827 | 0.140 |
| W | 563 | 13 | 0.40 | 0.14 | 225 | 3511 | 0.592 |
| X | 131 | 5 | 1.07 | 0.02 | 140 | 78 | 0.013 |
| Total | 1775 | 41 |  |  | 848 | 5927 | 1 |
| 2009 TV survey |  |  |  |  |  |  |  |
| U | 656 | 9 | 0.39 | 0.03 | 255 | 1476 | 0.174 |
| V | 425 | 6 | 0.60 | 0.08 | 255 | 2251 | 0.266 |
| W | 563 | 8 | 0.54 | 0.12 | 306 | 4644 | 0.549 |
| X | 131 | 3 | 1.17 | 0.02 | 153 | 93 | 0.011 |
| Total | 1775 | 26 |  |  | 969 | 8464 | 1 |

*Note: abundance estimates for these years based on figures prior to the 2009 revision of the dataseries. Differences between these figures and the revised figures shown on Table 3.5.8 are small.

Table 3.5.8. Nephrops, North Minch (FU11): Results of the 1994-2009 TV surveys.

| Year | Number of valid STATIONS | Mean DENSITY | Abundance | 95\% CONFIDENCE interval |
| :---: | :---: | :---: | :---: | :---: |
|  |  | burrows/m ${ }^{2}$ | millions | millions |
| 1994 | 41 | 0.38 | 665 | 99 |
| 1995 |  |  | urvey |  |
| 1996 | 38 | 0.25 | 439 | 62 |
| 1997 |  |  | urvey |  |
| 1998 | 38 | 0.41 | 728 | 103 |
| 1999 | 36 | 0.36 | 644 | 119 |
| 2000 | 39 | 0.53 | 946 | 109 |
| 2001 | 56 | 0.50 | 886 | 108 |
| 2002 | 37 | 0.61 | 1084 | 121 |
| 2003 | 41 | 0.80 | 1420 | 171 |
| 2004 | 38 | 0.80 | 1420 | 142 |
| 2005 | 41 | 0.70 | 1249 | 133 |
| 2006 | 30 | 0.81 | 1429 | 134 |
| 2007 | 36 | 0.55 | 978 | 122 |
| 2008 | 41 | 0.48 | 848 | 127 |
| 2009 | 26 | 0.55 | 969 | 184 |

Table 3.5.9. Nephrops, North Minch (FU11-13): Mean weight in the landings.

| YEAR | FU11 | FU12 | FU13 |
| :---: | :---: | :---: | :---: |
| 1990 | 21.31 | 19.90 | 24.21 |
| 1991 | 25.28 | 21.65 | 20.57 |
| 1992 | 21.58 | 24.01 | 25.08 |
| 1993 | 20.70 | 21.16 | 29.40 |
| 1994 | 23.38 | 24.88 | 19.14 |
| 1995 | 22.16 | 21.87 | 21.60 |
| 1996 | 26.63 | 23.02 | 24.14 |
| 1997 | 21.62 | 23.28 | 18.04 |
| 1998 | 23.57 | 22.09 | 16.74 |
| 1999 | 21.49 | 23.60 | 19.54 |
| 2000 | 23.17 | 24.81 | 19.06 |
| 2001 | 23.03 | 21.44 | 15.82 |
| 2002 | 22.86 | 23.60 | 18.59 |
| 2003 | 21.45 | 24.48 | 18.31 |
| 2004 | 23.62 | 24.02 | 17.46 |
| 2005 | 21.97 | 23.53 | 18.66 |
| 2006 | 21.68 | 23.15 | 18.53 |
| 2007 | 21.15 | 21.43 | 16.05 |
| 2008 | 22.18 | 21.41 | 18.10 |
| 2009 | 21.67 | 21.07 | 17.56 |
| Mean $(07-09)$ | 21.30 |  |  |

Table 3.5.10. Nephrops, North Minch (FU11): Adjusted TV survey abundance, landings, discard rate (proportion by number) and estimated harvest rate.

|  | Adjusted survey <br> (MiluIons) | Landings <br> (tonnes) | Discard rate (\%) | Harvest ratio* |
| :---: | :---: | :---: | :---: | :---: |
| 1999 | 484 | 3257 | 0.16 | 0.34 |
| 2000 | 711 | 3246 | 0.07 | 0.20 |
| 2001 | 666 | 3259 | 0.12 | 0.21 |
| 2002 | 815 | 3440 | 0.18 | 0.19 |
| 2003 | 1068 | 3268 | 0.19 | 0.14 |
| 2004 | 1068 | 3135 | 0.13 | 0.13 |
| 2005 | 939 | 2984 | 0.32 | 0.15 |
| 2006 | 1074 | 4160 | 0.30 | 0.21 |
| 2007 | 735 | 3968 | 0.07 | 0.24 |
| 2008 | 638 | 3799 | 0.11 | 0.27 |
| 2009 | 729 | 3497 | 0.20 | 0.22 |

*harvest rates previous to 2006 are unreliable.


Figure 3.5.1. Nephrops Functional Units in VIa and VIIa. North Minch (FU11), South Minch (FU12), Clyde (FU13), Irish Sea East (FU14) and Irish Sea West (FU15).


Figure 3.5.2. Nephrops in Division VIa. Landing (thousands tonnes) by FU and Other rectangles.


## Landings - International

LPUE - Scottish Nephrops trawlers


## Mean sizes - Scottish Nephrops trawlers



Figure 3.5.3. Nephrops, North Minch (FU11), Long-term landings, effort, lpue and mean sizes.


Figure 3.5.4. Nephrops, North Minch (FU11), Landings, effort and lpues by quarter and sex from Scottish Nephrops trawlers.


Figure 3.5.5. Nephrops, North Minch (FU11), Catch length frequency distribution and mean sizes (red line) for Nephrops in the North Minch, 1979-2009.

## Mean weight in landings



Figure 3.5.6. Nephrops, (FU11-13), individual mean weight in the landings from 1990-2009 (from Scottish market sampling data).


Figure 3.5.7. Nephrops, North Minch (FU11), TV survey station distribution and relative density (burrows $/ \mathrm{m}^{2}$ ), 2004-2009. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles in these figures are all scaled the same. Red crosses represent zero observations.


Figure 3.5.8. Nephrops, North Minch (FU11), Time-series of revised TV survey abundance estimates (not adjusted for bias), with 95\% confidence intervals, 1994-2009 (no survey 1995 and 1997).


Figure 3.5.9. Nephrops, North Minch (FU11), comparison of area of Nephrops ground defined by BGS sediment distribution (upper plot) and by distribution of VMS pings (shown in red) recorded from Nephrops trawlers $\mathbf{> 1 5} \mathbf{~ m}$ length (lower panel).

### 3.6 South Minch, FU12

## Type of assessment in 2010

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG and described in Section 2.2.

### 3.6.1 Ecosystem aspects

The South Minch Functional Unit 12 is located mid way down the west coast of Scotland (Figure 3.5.1).

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the South Minch Functional Unit these substrates are distributed according to prevailing hydrographic and baythmetric conditions. The area is characterised by numerous islands of varying size and sea lochs occur along the mainland coast. These topographical features create a diverse habitat with complex hydrography and a patchy distribution of soft sediments. A more continuous extensive area of sediment suitable for Nephrops occurs further offshore to the southwest. Figure 3.6.4 shows the distribution of sediment in the area.

Additional information on ecosystem aspects can be found in the Stock Annex.

### 3.6.2 The fishery in 2009

Two distinct fleets continued to operate in the South Minch during 2009, landing into the two main ports of Oban and Mallaig. Inshore, a large fleet of smaller vessels including creel boats operated throughout the year, whilst some larger twin riggers fished slightly further afield. $90 \%$ of boats are thought to fish for Nephrops at some time. Around 15 to 20 vessels are resident to Mallaig throughout the year. The local fleet has declined over the years. Approximately ten of these vessels are 'day boats', and approximately five are $17-19 \mathrm{~m}$ long twin riggers. Trips were typically of $1-3$ days usually operating within about two hours steaming distance.

Traditionally east coast vessels (mainly twin riggers from Fraserburgh) visit Mallaig in March or April, but in the last years there was a significant reduction in effort from visiting vessels. During the winter months, fishing activity is severely reduced in the South Minch due to the weather and small boats are often restricted to trawling in the sheltered sea-lochs. There is increasing overlap of the areas exploited by trawl and creel fishing (This is described further in the quality of assessment section illustrate the extent of trawling by some vessels). Boats on the west coast of Scotland are operating in accordance with the Scottish Conservation Credits Scheme and during 2009 were also required to fit 120 mm square meshed panels in accordance with the west coast emergency measures (Council Reg. (EU) 43/2009).

### 3.6.3 ICES advice for 2009 and 2010

ICES advice for 2009 based on Exploitation boundaries in relation to precautionary considerations was as follows:
"The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to the past two years (2006-2007). This corresponds to landings of no more than 5000 tonnes for the South Minch stock."

ICES advice for 2010 based on Single-stock exploitation boundaries was as follows:
"ICES advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Rate for Nephrops fisheries should not exceed $\mathrm{F}_{2008}$. This corresponds to landings of no more than 4126 t for the South Minch stock."

### 3.6.4 Management applicable in 2009 and 2010

Management applicable to this stock is included in management for Division VIa as a whole, and is described in Section 3.5.1.

### 3.6.5 Assessment

No specific concerns were raised in relation to the assessment method or data during the review of the 2009 assessment.

## Approach in 2010

The assessment in 2010 is based on a combination of examining trends in fishery indicators and underwater TV using an extensive data series for the South Minch FU12.

The assessment of Nephrops through the use of the UWTV survey data and other commercial fishery data follows the process defined by the benchmark WG and described in the Stock Annex.

The provision of advice in 2010 develops the process defined by the Benchmark WG. Section 2.2 outlines the WG approach to integrate WKFRAME recommendations in the provision of Fmsy proxies for Nephrops. The approach was developed based on intersessional work carried out by participants of the Benchmark and involving collaboration between WGNSSK and WGCSE.

Previous TV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. Creel fishing is important in the South Minch and increasingly operates across similar areas to the trawl fishery. For this reason the assessment is performed using combined length compositions from these fisheries.

## Data available

An overview of the data provided and used by the WG is shown in Table 2.1.

## Commerical catch and effort data

Official catch statistics (landings) reported to ICES are shown in Table 3.5.2. These relate to the whole of VIa of which the South Minch is a part. Landings for FU12 provided through national laboratories are presented in Table 3.6.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from Scotland, with low levels reported from the rest of the UK in the mid-1990s, and low levels more recently reported for Ireland. Total international reported landings in 2009 was 4282 tonnes, consisting of 3347 tonnes landed by trawlers and 900 tonnes landed by creel vessels. These estimates for total landings show a reduction from the high values in the previous two years to landings more typical of the late 1980s. The high landings of 2006-2008 are thought to have arisen through a combination of good recruitment in the mid-2000s feeding into the fished population, increased catching opportunities and to the introduction of the "buyers and sellers" regulations in the UK in 2006 which have increased the reliability of landings information. Landings
from creel vessels have remained relatively stable over the last four years, at close to 1000 tonnes, the highest level in the time-series.

Reported effort (given in days fished rather than hours since this is thought to be more reliable).by all Scottish Nephrops trawlers has fluctuated without trend in the most recent years after reaching a peak in the early 1990s. (Figures 3.6.1 and 3.6.2).

Sex ratio in the South Minch shows some variation but males consistently make the largest contribution to the annual landings. This occurs because males are available throughout the year and the fishery is also prosecuted in all quarters. Females on the other hand are mainly taken in the summer when they emerge after egg hatching (Figure 3.6.2).

Discarding of undersized and unwanted Nephrops occurs in this fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 2000. Discarding rates average around $21 \%$ by number in this FU (Table 3.6.5).

Studies (Guéguen, J. and Chareau, A., 1975; Sangster et al., 1997; Wileman et al., 1999) suggest that some Nephrops survive the discarding process, an estimate of $25 \%$ survival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard rate adjusted to account for some survival was estimated at the Benchmark Workshop (WKNEPH 2009) to be $16.7 \%$ (taking a three year average 2005-2007) and according to the agreed benchmark protocol this value is used in the provision of landings options for 2011.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Quarterly landings and discards-at-length data were available from Scotland and these sampling levels are shown in Table 3.5.4. Length compositions for the creel fishery are of landings only since the small numbers of discards survive well and are not considered to be removed from the population. Although assessments based on detailed catch analysis are not currently possible, examination of length compositions can provide a preliminary indication of exploitation effects.

Figure 3.6.3 shows a series of annual length frequency distributions for the period 1979 to 2009. Catch (removals) length compositions are shown for each sex along with the mean size for both. In both sexes the mean sizes have been fairly stable over time. Examination of the tails of the distributions above 35 mm (the length beyond which the effects of recruitment pulses and discarding are considered to be negligible) shows no evidence of reductions in relative numbers of larger animals.

The observation of relatively stable length compositions is further confirmed in the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings shown in Figure 3.6.1 and Table 3.6.2. This parameter might be expected to reduce in size if overexploitation were taking place but there is no evidence of this. The mean size of smaller animals ( $<35 \mathrm{~mm}$ ) in the catch (and landings) is also quite stable through time.

Mean weight in the landings is shown in Figure 3.5.6 and Table 3.5.9 and this also shows no systematic changes over the time-series.

## Natural mortality, maturity-at-age and other biological parameters

Biological parameter values are included in the Stock Annex.

## Research vessel data

Underwater TV surveys using a stratified random approach are available for this stock since 1995. Underwater television surveys of Nephrops burrow number and distribution, reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

The numbers of valid stations used in the final analysis in each year are shown in Table 3.6.4. On average, 35 stations have been considered valid each year, and then raised to a stock area of $5072 \mathrm{~km}^{2}$. In 2009 station numbers were the second lowest in the time-series, owing to time constraints on the RV survey and the presence of creels at a number of the planned station locations.

## Data analyses

## Exploratory analyses of survey data

Full details of the UWTV approach can be found in the Stock Annex and the Report of (WKNEPH) in 2009 (ICES, 2009).

A re-working of the UWTV survey abundance series for Division VIa was presented to the Nephrops Benchmark Workshop (WKNEPH) in 2009 (ICES, 2009) and further details of the technical changes to the camera can be found in the report of that workshop. The revised abundance estimates for FU12 from 1999 onwards were presented for the first time at WGCSE 2009 and are slightly higher than the previous values due to the field of view being smaller than previously calculated.

Table 3.6.3 shows the basic analysis for the three most recent TV surveys conducted in FU12. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. Due to the fact only one station was surveyed in the mud sediment type in 2008, it was not possible to calculate a sample variance for this area in the usual way. Instead an average of the three previous years was taken. Results in 2009 were typical of previous years.

Figure 3.6.4 shows the distribution of stations in recent TV surveys (2004-2009), with the size of the symbol reflecting the Nephrops burrow density. The most recent survey suggests higher abundance in the southeast part of the functional unit.. Table 3.6.4 and Figure 3.6 .5 show the time-series estimated abundance for the TV surveys, with $95 \%$ confidence intervals on annual estimates.

The review of the use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential biases were highlighted including those due to edge effects, species burrow misidentification and burrow occupancy. The cumulative bias correction factor estimated for FU12 was 1.32 meaning that the TV survey is likely to overestimate Nephrops abundance by $32 \%$.

## Final assessment

The underwater TV survey is presented as the best available information on the South Minch (FU12) Nephrops stock. This survey provides a fishery independent estimate of Nephrops abundance. The details of the 2009 survey is shown in Table 3.6.3 and compared with the 2007 and 2008 outcomes. At present it is not possible to extract any length or age structure information from the survey and it therefore only provides information on abundance over the area of the survey.

The 2009 TV survey data presented at this meeting shows that the abundance is more or less the same as in 2008, although the confidence limits are quite large around the estimate.

The TV survey results reported here do not cover the sea loch areas adjacent to the main South Minch grounds and should therefore be considered underestimates of the overall abundance. The sea lochs support an unknown but significant part of both the trawl and creel fishery. This issue is discussed further under quality of assessment.

### 3.6.6 Historic stock trends

The TV survey estimates of abundance for Nephrops in the South Minch show that the population has fluctuated without obvious trend over the period of the survey. The recently observed upturn gives an abundance which is just below the long-term average ( 2200 million animals). The bias adjusted abundance estimates from 1999-2009 (the period over which the survey estimates have been revised) is shown in Table 3.6.5. The stock is estimated to now be at 2035 million individuals as shown in Table 3.6.4.

Table 3.6.5 also shows the estimated harvest ratios over this period. These range from $7-27 \%$ over this period. (It is likely that prior to 2006, the harvest ratios are underestimates of the actual harvest ratios due to under-reporting of landings).

### 3.6.7 MSY considerations

A number of potential $\mathrm{F}_{\text {msy }}$ proxies are obtained from the per-recruit analysis for Nephrops and these are discussed further in Section 2.2 of this report. The analysis assumes the same input parameters (exploitation, discard ogive and biological parameters) as used at the Benchmark meeting in 2009. The complete range of the perrecruit $\mathrm{F}_{\text {msy }}$ proxies is given in the table below and the process for choosing an appropriate $F_{m s y}$ proxy is described in Section 2.2. Note that all $\mathrm{F}_{\mathrm{msy}}$ proxy harvest rate values are considered preliminary and may be modified following further data exploration and analysis.

For this FU, the absolute density observed on the UWTV survey is intermediate (average of just over $0.43 \mathrm{~m}^{-2}$ ) suggesting the stock has moderate productivity. In addition, the fishery in this area has been in existence since the 1960s and the population has been studied numerous times (Afonso-Dias, 1998; Howard and Hall, 1983). Historical harvest ratios in this FU have been variable but generally around the $\mathrm{F}_{35}$ \%SpR. The WG concluded that combined sex $\mathrm{F}_{35} \% \mathrm{SpR}$ is an appropriate F proxy for South Minch FU12 Nephrops. This is slightly below $\mathrm{F}_{\max }$ in males and is predicted to result in about $27 \%$ SPR for males; in excess of the $20 \%$ considered precautionary lower bound outlined in Section 2.2.

|  |  | Fbar(20-40 mm) |  |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fmult | M | F |  | M | F | T |
| F0.1 | M | 0.22 | 0.13 | 0.06 | 7.8 | 40.9 | 60.8 | 48.5 |
|  | F | 0.44 | 0.27 | 0.12 | 13.8 | 23.8 | 43.7 | 31.4 |
|  | T | 0.25 | 0.15 | 0.07 | 8.7 | 37.4 | 57.7 | 45.2 |
| $\mathrm{F}_{\text {max }}$ | M | 0.42 | 0.25 | 0.12 | 13.3 | 24.8 | 44.8 | 32.5 |
|  | F | 1.1 | 0.67 | 0.31 | 26.8 | 9.9 | 23.6 | 15.2 |
|  | T | 0.54 | 0.33 | 0.15 | 16.1 | 19.8 | 38.7 | 27.1 |
| $\mathrm{F}_{35 \% \mathrm{~S} \text { SR }}$ | M | 0.28 | 0.17 | 0.08 | 9.6 | 34.5 | 54.9 | 42.3 |
|  | F | 0.64 | 0.39 | 0.18 | 18.3 | 16.9 | 34.8 | 23.8 |
|  | T | 0.38 | 0.23 | 0.11 | 12.3 | 27.0 | 47.3 | 34.8 |

The $\mathrm{B}_{\text {trigger }}$ point for this FU (bias adjusted lowest observed UWTV abundance) is calculated as 1016 million individuals.

### 3.6.8 Landings forecasts

A landings prediction for 2011 was made for the South Minch (FU12) using the approach agreed at the Benchmark Workshop and outlined in the Section 2.2. The text table below shows landings predictions at various harvest ratios, including a selection of those equivalent to the per-recruit reference points discussed in Section 2 of this report and the harvest ratio in 2009 using the input parameters agreed at WKNEPH (ICES 2009). The landings prediction for 2011 at the $\mathrm{F}_{\text {msy }}$ proxy harvest ratio considered appropriate for the South Minch (i.e. $12.3 \%$ ) is 3809 tonnes. Since current harvest rate is above the $\mathrm{F}_{\text {msy }}$ proxy, the transition scheme towards the ICES MSY framework applies and would result in a landings estimate of 3995 tonnes for 2011.

The inputs to the landings forecast were as follows:
Mean weight in landings $(07-09)=23.8 \mathrm{~g}$
Discard rate $($ by number $)=15.7 \%$
Survey bias $=1.32$.

|  | Harvest rate | Survey Index (adjusted) | Implied fishery |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Retained number | Landings (tonnes) |
| $\mathrm{F}_{\mathrm{msy}}$ | 12.3\% | 1542 | 160 | 3809 |
| Fmsy transition | 12.9\% | 1542 | 168 | 3995 |
| $\mathrm{F}_{0.1(\mathrm{M})}$ | 7.8\% | 1542 | 101 | 2416 |
| $\mathrm{F}_{0.1(\mathrm{~T})}$ | 8.7\% | 1542 | 113 | 2694 |
| $\mathrm{F}_{35 \% \mathrm{SpR}(\mathrm{M})}$ | 9.6\% | 1542 | 125 | 2973 |
| $\mathrm{F}_{35 \% \mathrm{SpR}(\mathrm{T})}$ | 12.3\% | 1542 | 160 | 3809 |
| $\mathrm{F}_{2009}$ | 13.0\% | 1542 | 169 | 4026 |
| $\mathrm{F}_{\text {max }}(\mathrm{M})$ | 13.3\% | 1542 | 173 | 4119 |
| $\mathrm{Fmax}_{\text {(T) }}$ | 16.1\% | 1542 | 209 | 4986 |

$\mathrm{F}_{0.1(\mathrm{M}, \mathrm{T})}$ : Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the male or combined sex YPR curve.
$\mathrm{F}_{35 \% \mathrm{SPR}(\mathrm{M}, \mathrm{T}) \text { : Harvest ratio equivalent to fishing at a rate which results in male or com- }}$ bined SPR equal to $35 \%$ of the unfished level.
$F_{\max (M, T)}$ : Harvest ratio equivalent to fishing at a rate which maximises the male or combined YPR.

A discussion of $\mathrm{F}_{\mathrm{msy}}$ reference points for Nephrops is provided in Section 2.2.

### 3.6.9 Biological reference points

Precautionary approach biological reference points have not been determined for Nephrops stocks.

### 3.6.10 Quality of assessment and forecast

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in this fishery since 1990, and is considered to represent the trawl fishery adequately. In this assessment combined trawl and creel length compositions are used to account for the fact that the creel fishery accounts for over $20 \%$ of the landings, increasingly operates over similar areas to trawling, and exhibits a length composition composed of larger animals.

There are concerns over the accuracy of historical landings and effort data prior to 2006 when Buyers and Sellers was introduced and the reliability began to improve. Because of this the final assessment adopted is independent of official statistics. Incorporation of creel length compositions has also improved estimates of harvest ratios.

Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are on average greater during the most recent years, when abundance estimates have been slightly higher. The overlap of confidence intervals makes it difficult to determine which population changes are significant. Results suggest the population has fluctuated without trend.

There is a gap of 18 months between the survey and the start of the year for which the assessment is used to set management levels. It is assumed that the stock is in equilibrium during this period (i.e. recruitment and growth balance mortality) although this is rarely the case. The effect of this assumption on realised harvest rates has not been investigated.

The cumulative bias estimates for FU12 are largely based on expert opinion (See Annex). The precision of these bias corrections cannot yet be characterised.

The survey should be considered as a minimum estimate. Overall area of the ground is estimated by contoured sediment data. New VMS data linked to landings (from queries of the Scottish FIN database) suggest that not all areas are being considered in the current UWTV approach and as such, the absolute abundance estimate for this ground is likely to be an underestimate. Figure 3.6.6 illustrates differences between the British Geological Survey based sediment approach to estimating area and the activity of $>15 \mathrm{~m}$ trawlers- inclusion of smaller vessels would likely further modify this. Work is in progress to refine the area estimate.

The landings forecast for 2011 (equivalent to fishing at $\mathrm{F}_{35 \% \mathrm{SpR}}$ ) is 4009 tonnes. This is very close to the reported landings in 2009. In the provision of catch options based
on the absolute survey estimates additional uncertainties related to mean weight in the landings (which in this case are very stable) and the discard rates also arise. A three year average (2005-2007) of discard rate (adjusted to account for some survival of discarded animals) have been used in the calculation of catch options. The recent observed discard rate has however shown a 50\% decline in 2009.

### 3.6.11 Status of the stock

The UWTV survey indicates that the population declined from a record high in 2004 to record low in 2007 but is at a higher level again in 2009. The stable mean sizes in the length compositions of catches (of individuals $>35 \mathrm{~mm} \mathrm{CL}$ ) and recent fall in estimated harvest ratios (removals/TV abundance) to the equivalent of the FMSY proxy suggests that the stock is now being exploited sustainably.

### 3.6.12 Management considerations

The WG, ACFM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could confer controls to ensure effort and catch were in line with resources available.

Creel fishing takes place in this area but overall effort in terms of creel numbers is not known and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the South Minch and STECF continues to estimate that discards of whiting and haddock are high in VIa generally. It is important that efforts are made to ensure that unwanted bycatch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod under the Scottish Conservation Credits scheme and the West of Scotland emergency measures (Council Reg. (EU) 43/2009), include the implementation of larger meshed square meshed panels ( 120 mm ) and real time closures to avoid cod.

The implementation of buyers and sellers legislation in the UK in 2006 is improving the reliability of fishery statistics but the transition period is accompanied in some cases by large changes in landings which produce significant changes in the lpue and cpue series that cannot be completely attributed to changes in stock. Until a sufficient time-series of reliable data has built up, use of fishery catch and effort data in the assessment process should be avoided.

### 3.6.13 References

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Table 3.6.1. Nephrops, South Minch (FU12), Nominal Landings of Nephrops, 1981-2009, as officially reported.

| Year | UK Scotland |  |  |  | Other UK | Ireland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops trawl | Other trawl | Creel | Sub-total |  |  |  |
| 1981 | 2965 | 254 | 432 | 3651 | 0 | 0 | 3651 |
| 1982 | 2925 | 207 | 420 | 3552 | 0 | 0 | 3552 |
| 1983 | 2595 | 361 | 456 | 3412 | 0 | 0 | 3412 |
| 1984 | 3228 | 478 | 594 | 4300 | 0 | 0 | 4300 |
| 1985 | 3096 | 424 | 488 | 4008 | 0 | 0 | 4008 |
| 1986 | 2694 | 288 | 502 | 3484 | 0 | 0 | 3484 |
| 1987 | 2927 | 418 | 546 | 3891 | 0 | 0 | 3891 |
| 1988 | 3544 | 364 | 555 | 4463 | 10 | 0 | 4473 |
| 1989 | 3846 | 338 | 561 | 4745 | 0 | 0 | 4745 |
| 1990 | 3732 | 262 | 436 | 4430 | 0 | 0 | 4430 |
| 1991 | 3597 | 341 | 503 | 4441 | 1 | 0 | 4442 |
| 1992 | 3479 | 208 | 549 | 4236 | 1 | 0 | 4237 |
| 1993 | 3608 | 193 | 649 | 4450 | 5 | 0 | 4455 |
| 1994 | 3743 | 265 | 404 | 4412 | 3 | 0 | 4415 |
| 1995 | 3442 | 716 | 508 | 4666 | 14 | 0 | 4680 |
| 1996 | 3107 | 419 | 468 | 3994 | 1 | 0 | 3995 |
| 1997 | 3519 | 331 | 492 | 4342 | 3 | 1 | 4345 |
| 1998 | 2851 | 340 | 538 | 3729 | 0 | 0 | 3730 |
| 1999 | 3165 | 359 | 513 | 4037 | 0 | 14 | 4051 |
| 2000 | 2939 | 312 | 699 | 3950 | 0 | 2 | 3952 |
| 2001 | 2823 | 393 | 767 | 3983 | 0 | 9 | 3992 |
| 2002 | 2234 | 315 | 742 | 3291 | 0 | 14 | 3305 |
| 2003 | 2812 | 203 | 858 | 3873 | 0 | 6 | 3879 |
| 2004 | 2865 | 104 | 880 | 3849 | 0 | 19 | 3868 |
| 2005 | 2810 | 46 | 953 | 3809 | 1 | 31 | 3841 |
| 2006 | 3569 | 19 | 922 | 4510 | 9 | 35 | 4554 |
| 2007 | 4436 | 8 | 958 | 5402 | 19 | 30 | 5451 |
| 2008 | 4432 | 5 | 895 | 5332 | 2 | 13 | 5347 |
| 2009 | 3347 | 20 | 900 | 4267 | 4 | 11 | 4282 |

[^4]Table 3.6.2. Nephrops, South Minch (FU12): Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish catches and landings, 1981-2009.

| Year | Catches |  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | < 35 mm CL |  | $>35 \mathrm{~mm} \mathrm{CL}$ |  |
|  | Males | Females | Males | Females | Males | Females |
| 1981 | 28.2 | 26.4 | 29.6 | 27.5 | 41.5 | 38.0 |
| 1982 | 27.8 | 27.1 | 28.7 | 28.8 | 41.7 | 41.3 |
| 1983 | 28.6 | 26.5 | 29.3 | 27.6 | 39.5 | 37.6 |
| 1984 | 27.9 | 26.3 | 28.4 | 27.0 | 39.8 | 38.0 |
| 1985 | 27.9 | 27.5 | 28.6 | 28.5 | 40.0 | 37.6 |
| 1986 | 28.4 | 27.9 | 29.3 | 28.9 | 39.5 | 37.3 |
| 1987 | 28.3 | 26.6 | 29.2 | 28.1 | 39.8 | 37.6 |
| 1988 | 29.3 | 27.7 | 30.4 | 29.7 | 39.5 | 38.6 |
| 1989 | 28.6 | 28.1 | 29.8 | 29.4 | 39.5 | 38.4 |
| 1990 | 28.0 | 27.5 | 29.3 | 29.0 | 39.4 | 38.5 |
| 1991 | 29.4 | 27.5 | 29.9 | 27.9 | 39.0 | 38.5 |
| 1992 | 29.6 | 28.6 | 31.0 | 29.8 | 39.5 | 38.0 |
| 1993 | 29.0 | 27.8 | 30.0 | 28.5 | 39.5 | 38.0 |
| 1994 | 29.8 | 28.0 | 30.8 | 29.2 | 39.3 | 38.1 |
| 1995 | 29.5 | 28.2 | 30.0 | 28.4 | 39.4 | 38.0 |
| 1996 | 28.9 | 28.5 | 30.4 | 29.8 | 39.9 | 38.1 |
| 1997 | 29.3 | 28.7 | 30.6 | 29.6 | 39.8 | 37.8 |
| 1998 | 28.6 | 27.6 | 30.4 | 28.7 | 39.1 | 38.0 |
| 1999 | 28.6 | 27.7 | 30.0 | 29.5 | 39.4 | 38.3 |
| 2000 | 28.9 | 28.3 | 30.9 | 30.0 | 39.7 | 38.5 |
| 2001 | 27.7 | 27.3 | 29.7 | 28.8 | 39.6 | 38.1 |
| 2002 | 29.1 | 27.8 | 30.4 | 29.0 | 39.5 | 38.8 |
| 2003 | 29.0 | 28.1 | 30.4 | 29.5 | 39.8 | 38.4 |
| 2004 | 28.8 | 28.1 | 30.1 | 29.8 | 39.5 | 38.8 |
| 2005 | 28.1 | 27.8 | 30.4 | 29.5 | 39.8 | 38.6 |
| 2006 | 29.2 | 28.0 | 30.5 | 28.8 | 39.5 | 38.1 |
| 2007 | 29.7 | 28.2 | 29.9 | 28.2 | 40.0 | 38.3 |
| 2008 | 28.6 | 27.5 | 29.4 | 28.5 | 39.6 | 38.1 |
| 2009 | 28.7 | 27.8 | 29.7 | 28.6 | 40.0 | 38.3 |

[^5]Table 3.6.3. Nephrops South Minch (FU12). Results by stratum of the 2007-2009 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM - Sandy mud, MS Muddy sand).

|  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 TV Survey |  |  |  |  |  |  |  |
| M | 303 | 3 | 0.21 | 0.01 | 65 | 372 | 0.008 |
| SM | 2741 | 15 | 0.30 | 0.07 | 822 | 33184 | 0.727 |
| MS | 2028 | 21 | 0.24 | 0.06 | 490 | 12092 | 0.265 |
| Total | 5072 | 39 |  |  | 1377* | 45647 | 1 |
| 2008 TV Survey |  |  |  |  |  |  |  |
| M | 303 | 1 | 0.58 | 0.05 | 176 | 4593 | 0.037 |
| SM | 2741 | 18 | 0.45 | 0.19 | 1227 | 78145 | 0.636 |
| MS | 2028 | 14 | 0.36 | 0.14 | 718 | 40157 | 0.327 |
| Total | 5072 | 33 |  |  | 2121* | 122895 | 1 |
| 2009 TV Survey |  |  |  |  |  |  |  |
| M | 303 | 2 | 0.135 | 0.004 | 41 | 186 | 0.001 |
| SM | 2741 | 13 | 0.447 | 0.207 | 906 | 65406 | 0.373 |
| MS | 2028 | 10 | 0.397 | 0.146 | 1088 | 109660 | 0.626 |
| Total | 5072 | 25 |  |  | 2035 | 175252 | 1 |

*Note: abundance estimates for these years based on figures prior to the 2009 revision of the dataseries. Differences between these figures and the revised figures shown on Table 3.6.4 are small.

Table 3.6.4. Nephrops, South Minch (FU12): Results of the 1995-2009 TV surveys.

|  | Stations | Mean density | Abundance | 95\% <br> confidence interval |
| :---: | :---: | :---: | :---: | :---: |
| Year |  | burrows/m² | millions | millions |
| 1995 | 33 | 0.30 | 1520 | 331 |
| 1996 | 21 | 0.38 | 1945 | 700 |
| 1997 | 36 | 0.28 | 1434 | 244 |
| 1998 | 38 | 0.38 | 1916 | 306 |
| 1999 | 37 | 0.28 | 1433 | 343 |
| 2000 | 41 | 0.48 | 2447 | 460 |
| 2001 | 47 | 0.53 | 2689 | 606 |
| 2002 | 31 | 0.49 | 2507 | 749 |
| 2003 | 25 | 0.56 | 2847 | 998 |
| 2004 | 38 | 0.67 | 3377 | 625 |
| 2005 | 33 | 0.57 | 2914 | 977 |
| 2006 | 36 | 0.48 | 2436 | 789 |
| 2007 | 39 | 0.26 | 1341 | 205 |
| 2008 | 33 | 0.42 | 2123 | 548 |
| 2009 | 25 | 0.40 | 2035 | 837 |

Table 3.6.5. Nephrops, South Minch (FU12): Adjusted TV survey abundance, landings, discard rate proportion by number) and estimated harvest rate.

|  | Adjusted <br> survey <br> (millions) | Landings <br> (tonnes) | Discard rate (\%) | Harvest ratio* |
| :---: | :---: | :---: | :---: | :---: |
| 1999 | 1086 | 4051 | 0.15 | 0.16 |
| 2000 | 1854 | 3952 | 0.19 | 0.09 |
| 2001 | 2037 | 3992 | 0.28 | 0.11 |
| 2002 | 1899 | 3305 | 0.18 | 0.07 |
| 2003 | 2157 | 3879 | 0.21 | 0.08 |
| 2004 | 2558 | 3868 | 0.24 | 0.07 |
| 2005 | 2208 | 3841 | 0.26 | 0.08 |
| 2006 | 1845 | 4554 | 0.14 | 0.11 |
| 2007 | 1016 | 5451 | 0.22 | 0.27 |
| 2008 | 1608 | 5347 | 0.25 | 0.17 |
| 2009 | 1542 | 4282 | 0.12 | 0.13 |

[^6]
## Landings - International



Effort - Scottish Nephrops trawlers


LPUE - Scottish Nephrops trawlers


Mean sizes - Scottish Nephrops trawlers


Figure 3.6.1. Nephrops, South Minch (FU12), Long-term landings, effort, lpue and mean sizes.


Figure 3.6.2. Nephrops, South Minch (FU12), Landings, effort and lpues by quarter and sex from Scottish Nephrops trawlers.


Figure 3.6.3. Nephrops. South Minch (FU12). Catch length frequency distribution and mean sizes (red line) for Nephrops in the South Minch, 1979-2009.


Figure 3.6.4. Nephrops, South Minch (FU12), TV survey station distribution and relative density (burrows $/ \mathrm{m}^{2}$ ), 2004-2009. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles in this figure are all scaled the same. Red crosses represent zero observations.


Figure 3.6.5. Nephrops, South Minch (FU12), Time-series of revised TV survey abundance estimate (not adjusted for bias), with 95\% confidence intervals, 1995-2009.


Figure 3.6.6. Nephrops, South Minch (FU12), comparison of area of Nephrops ground defined by BGS sediment distribution (upper plot) and by distribution of VMS pings (shown in red) recorded from Nephrops trawlers >15 m length (lower panel).

### 3.7 Clyde, FU13

## Type of assessment in 2010

The assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG and described in Section 2.2.

### 3.7.1 Ecosystem aspects

The Clyde FU comprises two distinct patches in the Firth of Clyde and the Sound of Jura, to the east and west of the Mull of Kintyre respectively. The hydrography of the two subareas differs with the Sound of Jura characterised by stronger tidal currents and the Firth of Clyde exhibiting features of a lower energy environment with a shallow entrance sill.

Owing to its burrowing behaviour, the distribution of Nephrops is restricted to areas of mud, sandy mud and muddy sand. Within the two patches these substrates are distributed according to prevailing hydrographic and baythmetric conditions. The available area of suitable sediment is smaller in the Sound of Jura, occupying only the deepest parts of the Sound, while in the Firth of Clyde these sediments predominate.

Additional information on ecosystem aspects can now be found in the Stock Annex.

### 3.7.2 The fishery in 2009

Around 35 Trawlers ranging from 9.9 m to 20 m operated in the Clyde during 2009. Vessels were all using 80 mm codends with 120 mm minimum square mesh panels, in line with west coast emergency measures conditions (Council Reg. (EU) 43/2009). The most significant landings were made at the main Clyde landing ports of Troon, Girvan, Largs on the East side of the Clyde and Campbelltown, Tarbert, and Carradale on the west side of the Clyde. Almost all of the Clyde Nephrops fleet fish daily trips. Vessels in the Clyde tend to stick the same gear type but traditionally some will swap between Nephrops and scallop gear during the year. Fishing in the Clyde was generally steady through the year although there is a dip in catches during April and May. At the end of the summer, a large number of local skippers complained about the large number of jellyfish in the nets. In common with other years a small bycatch of fish was taken in the Clyde consisting mainly of cod, hake and whiting.

A few Northern Irish boats fish the Clyde at varying times of the year according to weather and catch rates. These boats fish mainly for tails, landing into Campbeltown or Troon.

Mobile gear is banned in the Inshore Clyde from Friday night to Sunday night as are vessels greater than 21 m in length. An increasing number of creel boats operate in the Clyde. Creeling activity now takes place quite widely in the northern parts of the Firth operating on some of the same grounds but often taking place during the weekend trawling ban. Only about a third of creelers operated throughout the year, the rest prosecuted a summer fishery.

There were numerous problems for the fleet during 2009, including poor catches, poor prices, high fuel prices and lack of crew. The recession played a part with the markets and this meant poor prices for catches. The tail market was the most affected with buyers clearing out frozen stock before restocking.

### 3.7.3 ICES advice for 2009 and 2010

The ICES conclusions in 2009 in relation to State of the Stock were as follows:
"The stock is being exploited unsustainably. The current harvest rate is well above $\mathrm{F}_{\text {max. }}$. The UWTV survey indicates that the population has been at a relatively high level since 2003 except for 2007."

The ICES advice for 2009 (Exploitation boundaries in relation to precautionary considerations):
"The current fishery appears sustainable. Therefore, ICES recommends that the Nephrops fisheries should not be allowed to increase relative to the past two years (2006-2007). This corresponds to landings of no more than 5700 tonnes for the Firth of Clyde stock."

The ICES advice for 2010 (Single-stock exploitation boundaries) was as follows:
"ICES advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Rate for Nephrops fisheries should not exceed $\mathrm{F}_{\text {max. }}$. This corresponds to landings of no more than 3855 t for the Firth of Clyde stock."

### 3.7.4 Management applicaple to 2009 and 2010

Management is at the ICES subarea level as described at the beginning of Section 3.5. In 2009, ICES again reiterated its advice that Nephrops stocks should be managed at the FU level.

### 3.7.5 Assessment

The Review of the 2009 assessment concluded as follows:
"RG agrees with the WG on the assessment and feels it follows the protocol described in the Stock Annex. The short-term projection gives various harvest rates and this should be used to assign the TAC. The idea of fishing at a level above $F_{\max }$ is unsettling and should be avoided especially for a stock that utilizes such a basic assessment."

## Approach in 2010

The assessment in 2010 is based on a combination of examining trends in fishery indicators and underwater TV using an extensive dataseries for the Firth of Clyde component of FU13. For the first time an attempt is also made to use the more limited UWTV data available for the Sound of Jura subarea.

The assessment of Nephrops through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG and described in Section 2.2.

The provision of advice in 2010 develops the process defined by the Benchmark WG and described in Section 2.2 and attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES in 2010. Intersessional work carried out by participants of the Benchmark and involving collaboration between WGNSSK and WGCSE is described in the working papers, etc.

Previous TV based assessments have derived predicted landings by applying a harvest rate approach to populations described in terms of length compositions from the trawl component of the fishery. In recent years, creel fishing has become more impor-
tant in the Firth of Clyde and operates across similar areas to the trawl fishery. For this reason the assessment is performed using combined length compositions.

## Data available

An overview of the data provided and used by the WG is shown in Table 2.1.

## Commercial catch and effort data

Official catch statistics (landings) reported to ICES are shown in Table 3.7.1. These relate to the whole of VIa of which the Clyde FU is a part. Landings statistics for FU13 provided through national laboratories are presented in Table 3.7.1, broken down by country and by gear type. Landings from this fishery are predominantly reported from Scotland, although the remainder of the UK also contributed about $6 \%$ in 2009; landings from Northern Ireland form the main part of this. Total international reported landings decreased markedly in 2009 but remain well above the average for the time-series ( $\sim 3712$ tonnes), and consisted of 4303 tonnes landed by trawlers (Scottish and other UK) and 190 tonnes landed by creel vessels. Creel landings have increased in the most recent years but remain at a low level compared to other methods and to the creel fisheries elsewhere on the west coast of Scotland.

Table 3.7.2 shows the split in landings between the two subareas comprising FU13. Most of the landings are presently taken from the Firth of Clyde subarea with only about $2 \%$ from the Sound of Jura. Earlier in the time-series the Sound of Jura contributed as much a $20 \%$. The decline has occurred through a progressive reduction in fishing activity in the area. The main reason for this is probably related to the size composition in the population which is characterised by small Nephrops (Bailey and Chapman, 1983) whereas the market has increasingly favoured larger whole animals.

The introduction of the "buyers and sellers" regulation in the UK in 2006 has led to increased reliability in the reported landings.

Uncertainities over the accuracy of the effort data emerged just prior to the WG. In an effort to improve reliability, effort was extracted and expressed in terms of days fished (since the logbook field for hours is not mandatory). Preliminary examination of the new effort-series showed a marked discontinuity around 1995 with a large and inexplicable drop in effort in days. Further investigation revealed that at this time the process of recording days effort in the split rectangle region of the Clyde changed. This will require some additional work to establish if a reliable series can be reinstated. For the present, long-term trends in effort and lpue/cpue are not reported here. It is not thought however, that the change has affected the intra-annual, quarterly patterns of effort and lpue and these have been included.

Sex ratio in the Firth of Clyde shows some variation but males consistently make the largest contribution to the annual landings. This occurs because males are available throughout the year and the fishery is also prosecuted in all quarters. Females on the other hand are mainly taken in the summer when they emerge after egg hatching. (Figure 3.7.2).

Discarding of undersized and unwanted Nephrops occurs in the Firth of Clyde fishery, and quarterly discard sampling has been conducted on the Scottish Nephrops trawler fleet since 2000. Discarding rates are high in this FU and average around $31 \%$ by number in this FU since 1999. In 2009, discard rates were estimated to be higher than average at $39 \%$ by number (Table 3.7.8).

Studies (Guéguen, J. and Chareau, A., 1975; Sangster et al., 1997; Wileman et al., 1999) suggest that some Nephrops survive the discarding process, an estimate of $25 \%$ sur-
vival is assumed for this FU in order to calculate removals (landings + dead discards) from the population. The discard rate adjusted to account for some survival was estimated at the benchmark workshop to be $18.6 \%$ (taking a 3 year average 2005-2007) and according to the agreed benchmark protocol this value is used in the provision of landings options for 2011. This relatively low figure is due to a large drop in discarding in 2006, possibly as a result of reduced recruitment in this year that led to the low TV survey abundance estimate in 2007.

## Length compositions

Length compositions of landings and discards are obtained during monthly market sampling and quarterly on-board observer sampling respectively. Quarterly landings and discards-at-length data were available for the Firth of Clyde from Scotland and these sampling levels are shown in Table 3.5.4. Length compositions for the creel fishery are of landings only since the small numbers of discards survive well and are not considered to be removed from the population. Sampling of length compositions in the Sound of Jura is more infrequent and only limited data are available. Although assessments based on detailed catch analysis are not presently considered advisable, examination of length compositions can provide a preliminary indication of exploitation effects.

Figure 3.7.3 shows a series of annual Firth of Clyde length frequency distributions for the period 1979 to 2009. Catch (removals) length compositions are shown for each sex along with the mean size for both. In both sexes the mean sizes have been fairly stable over time. Examination of the tails of the distributions above 35 mm (the length beyond which the effects of recruitment pulses and discarding are considered to be negligible) shows no evidence of reductions in relative numbers of larger animals.

The observation of relatively stable length compositions is further confirmed in the series of mean sizes of larger Nephrops ( $>35 \mathrm{~mm}$ ) in the landings shown in Figure 3.7.1 and Table 3.7.3. This parameter might be expected to reduce in size if overexploitation were taking place but there is no evidence of this. The mean size of smaller animals ( $<35 \mathrm{~mm}$ ) in the catch (and landings) is also quite stable through time, although in the most recent year the mean size of females in the catch below 35 mm has decreased quite markedly, suggesting possible good recruitment.

Mean weight in the Firth of Clyde landings is shown in Figure 3.5.6 and Table 3.5.9 and this also shows no systematic changes over the time-series.

## Natural mortality, maturity at age and other biological parameters

Biological parameter values are included in the Stock Annex.

## Research vessel data

Underwater TV surveys are available for both subareas since 1995 although the Sound of Jura has been sampled more infrequently. Underwater television surveys of Nephrops burrow number and distribution reduce the problems associated with traditional trawl surveys that arise from variability in burrow emergence of Nephrops. TV surveys are targeted at known areas of mud, sandy mud and muddy sand in which Nephrops construct burrows.

The UWTV in the Firth of Clyde subarea is carried out using a stratified random approach. The numbers of valid stations used in the final analysis in each year are shown in Table 3.7.4. On average, 37 stations have been considered valid each year,
and then raised to the estimated area of the ground available for Nephrops- $2080 \mathrm{~km}^{2}$ based on contoured superficial sediment information (British Geological Surveys).

The number of valid stations in the Sound of Jura is shown in Table 3.7.6.

## Data analyses

## Exploratory analyses of survey data

Full details of the UWTV approach can be found in the Stock Annex and the Report of (WKNEPH) in 2009 (ICES, 2009).

A re-working of the UWTV survey abundance series for Division VIa was presented to the Nephrops Benchmark Workshop (WKNEPH) in 2009 (ICES, 2009) and further details of the technical changes to the camera can be found in the Report of that workshop. The revised abundance estimates for FU13 from 1999 onwards were presented for the first time at WGCSE 2009 and are slightly higher than the previous values due to the field of view being smaller than previously calculated.

Table 3.7.4 shows the basic analysis for the most recent TV surveys conducted in the Firth of Clyde. The table includes estimates of abundance and variability in each of the strata adopted in the stratified random approach. The areas of all sediment types (mud, muddy sand and sandy mud) in this region are very similar and as such the number of stations surveyed in each sediment type is also similar. Basic analysis for the Sound of Jura is shown in Table 3.7.6.

Figure 3.7.4 shows the distribution of stations in recent TV surveys (2004-2009) across FU13 (the two distinct subareas can be clearly seen) with the size of the symbols reflecting the Nephrops burrow density. Table 3.7.5 and Figure 3.7.5 show the timeseries estimated abundance for the TV surveys in the Firth of Clyde, with $95 \%$ confidence intervals on annual estimates. Similar information for the Sound of Jura is shown in Table 3.7.7 and Figure 3.7.6.

The use of the UWTV surveys for Nephrops in the provision of advice was extensively reviewed by WKNEPH (ICES, 2009). A number of potential biases were highlighted including those due to edge effects, species burrow mis-identification and burrow occupancy. The cumulative bias correction factor estimated for the Firth of Clyde was 1.19 meaning that the TV survey is likely to overestimate Nephrops abundance by $19 \%$. A review of the Sound of Jura biases has not so far been carried out; biases are here assumed to be similar to the Firth of Clyde.

## Final Assessment

The underwater TV surveys are presented as the best available information on the stocks of Nephrops in the two subareas of FU13. The surveys provide fishery independent estimates of Nephrops abundance. The details of the 2009 Firth of Clyde survey are shown in Table 3.7.4 and compared with the 2007 and 2008 outcome. The details of the 2009 Sound of Jura survey are shown in Table 3.7.6. At present it is not possible to extract any length or age structure information from the survey and it therefore only provides information on abundance over the area of the survey.

The 2009 TV survey data presented at this meeting shows that the abundance in the Firth of Clyde has dropped slightly but remains at the upper end of the values observed throughout the time-series. Confidence limits are quite high for this stock.

The 2009 TV survey data presented at this meeting shows that the abundance in the Sound of Jura is similar to the previous estimate in 2007.

The TV survey results reported here do not cover the sea loch areas adjacent to the main Firth of Clyde and Sound of Jura areas and should therefore be considered underestimates of the overall biomass. This issue is discussed further under quality of assessment.

### 3.7.6 Historic stock trends

The TV survey estimates of abundance for Nephrops in the Firth of Clyde suggest that the population increased until the mid-2000s implying a sustained period of increased recruitment. Following this, abundance has declined and fluctuated around the values previously observed in the early 2000s just prior to the maximum. The bias adjusted abundance estimates from 1999-2009 (the period over which the survey estimates have been revised) is shown in Table 3.7.8. The latest bias adjusted stock estimate is 1499 million individuals.

Table 3.7.8 also shows the estimated harvest ratios over this period. These range from $12-51 \%$ over this period. (It is unlikely that prior to 2006, the estimated harvest ratios are representative of actual harvest ratios due to under-reporting of landings).

Results for the Sound Jura are sparser and are associated with large confidence intervals particularly in 2002 and 2006. Table 3.7.9 summarises the bias adjusted estimates of abundance and harvest rates where available.

### 3.7.7 MSY considerations

A number of potential $\mathrm{F}_{\mathrm{msy}}$ proxies are obtained from the per-recruit analysis for Nephrops and these are discussed further in Section 2.2 of this report. The analysis assumes the same input parameters (exploitation, discard ogive and biological parameters) as was used at the Benchmark meeting in 2009. The complete range of the per-recruit $\mathrm{F}_{\text {msy }}$ proxies for the Firth of Clyde sub area is given in the table below and the process for choosing an appropriate $\mathrm{F}_{\text {msy }}$ proxy is described in Section 2.2. Note that all $\mathrm{F}_{\text {msy }}$ proxy harvest rate values are considered preliminary and may be modified following further data exploration and analysis.

For the Firth of Clyde subarea of this FU, the absolute density observed on the UWTV survey is generally high (average of over $0.8 \mathrm{~m}^{-2}$ for entire series and around $1.0 \mathrm{~m}^{-2}$ for the last five years suggesting the stock has relatively high productivity. In addition, the fishery in this area has been in existence since the 1960s and the population and biological parameters have been studied numerous times (Bailey and Chapman, 1983; Tuck et al., 1997; Tuck et al., 1999). Historical harvest ratios in this FU have been generally high at or above $\mathrm{F}_{\text {max. }}$. An appropriate $\mathrm{F}_{\text {msy }}$ proxy is considered therefore to be the total population $\mathbf{F}_{\text {max }}$ which is predicted to deliver an $\mathbf{F}_{35 \% \mathrm{SpR}}$ of about $\mathbf{2 2 \%}$ for males; considered precautionary for this species (See Section 2.2).

|  |  | Fmult | Fbar(20-40 mm) |  | HR (\%) | SPR (\%) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | M | F |  | M | F | T |
| $\mathrm{F}_{0.1}$ | M | 0.17 | 0.15 | 0.06 | 8.7 | 40.2 | 66.8 | 49.1 |
|  | F | 0.43 | $0.37$ | 0.14 | 21.1 | 16.2 | 40.7 | 24.4 |
|  | T | 0.19 | 0.16 | 0.06 | 9.7 | 36.9 | 64.0 | 45.9 |
| $\mathrm{F}_{\text {max }}$ | M | 0.27 | 0.23 | 0.09 | 13.6 | 27.0 | 54.4 | 36.2 |
|  | F | 0.71 | 0.61 | 0.24 | 34.0 | 8.3 | 26.5 | 14.3 |
|  | T | $0.33$ | 0.28 | $0.11$ | 16.4 | 21.9 | 48.6 | 30.8 |
| $\mathrm{F}_{35 \% \mathrm{SPR}}$ | M | 0.21 | 0.18 | 0.07 | 10.7 | 34.0 | 61.4 | 43.1 |
|  | F | 0.53 | 0.46 | 0.18 | 25.7 | 12.4 | 34.6 | 19.8 |
|  | T | 0.29 | 0.25 | 0.10 | 14.5 | 25.1 | 52.4 | 34.2 |

The $B_{\text {trigger }}$ point for this FU (bias adjusted lowest observed UWTV abundance) is calculated as 579 million individuals.

Yield-per-recruit analysis is not yet available for the Sound of Jura subarea of this FU and so proxies from the Firth of Clyde (shown in the table above) are used to provide a first approach. The absolute density observed on the UWTV survey is generally high (average of about $0.8 \mathrm{~m}^{-2}$ over the time-series and around $1 \mathrm{~m}^{-2}$ over the last five years) suggesting the stock has relatively high productivity. A number of studies have investigated biology and the area is acknowledged as having high abundance for many years. However, the time-series of TV data is more fragmented and sampling is at a relatively low level; confidence intervals are larger. The fishery in this area has been in existence since the 1960s but in recent times has operated at a low level and harvest ratios in this FU have been low An appropriate $\mathrm{F}_{\text {msy }}$ proxy is considered therefore to be the total population $\mathrm{F}_{35 \% \mathrm{SpR}}$ which is predicted to deliver an $\mathbf{F}_{35 \% \mathrm{SpR}}$ of about $\mathbf{2 5 \%}$ for males; above the level considered precautionary for this species (See Section 2.2).

The $B_{\text {trigger }}$ point for this FU (bias adjusted lowest observed UWTV abundance) has not been defined but is expected to be below 200 million individuals.

### 3.7.8 Landings forecasts

Landings prediction for 2011 were made for the Firth of Clyde and Sound of Jura subareas of the Clyde FU13 using the approach agreed at WKNEPH 2009 and outlined in the Section 2.2 The tables below shows landings predictions at various harvest ratios, including a selection of those equivalent to the per-recruit reference points discussed in Section 2 of this report and the harvest ratio in 2009 using the input parameters agreed at WKNEPH (ICES 2009). The landings prediction for 2011 at the $\mathrm{F}_{\mathrm{msy}}$ proxy harvest ratio considered appropriate for the Firth of Clyde (i.e. $16.4 \%$ ) is 2804 tonnes. There is a transition stage as the current harvest ratio is above the $\mathrm{F}_{\mathrm{msy}}$ proxy in 2011 this gives landings of 4121 t .

For the Sound of Jura subarea, the landings prediction for 2011 at the Fmsy proxy harvest ratio of $14.5 \%$ is 515 t . There is no transition stage since the current position is below the $\mathrm{F}_{\text {msy }}$ proxy.

The inputs to the landings forecast for the Firth of Clyde and Sound of Jura were as follows:

Mean weight in landings in Firth of Clyde (07-09) $=17.8 \mathrm{~g}$
Mean weight in landings in Sound of Jura (07-09) $=22.1 \mathrm{~g}$

Discard rate (by number) $=36.0 \%$
Survey bias $=1.19$ (as calculated at WKNEPH 2009).

Firth of Clyde

|  | Harvest rate | Survey Index <br> (adjusted) | Implied fishery |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Retained number | Landings (tonnes) |
| $\mathrm{F}_{\mathrm{msy}}$ | 16.4\% | $1499$ | 157 | 2804 |
| $\mathrm{F}_{\text {msy transition }}$ | $24.1 \%$ | $1499$ | $231$ | 4121 |
| $\mathrm{F}_{0.1(\mathrm{M})}$ | 8.7\% | 1499 | 83 | 1488 |
| F0.1(T) | $9.7 \%$ | $1499$ | $93$ | 1659 |
| $\mathrm{F}_{35 \% \mathrm{SpR}(\mathrm{M})}$ | $10.7 \%$ | $1499$ | $103$ | $1830$ |
| $\mathrm{F}_{\max (\mathbb{M})}$ | $13.6 \%$ | $1499$ | $130$ | $2325$ |
| $\mathrm{F}_{35 \%}$ \% SRR(T) | 14.5\% | 1499 | $139$ | 2479 |
| $\mathrm{F}_{\text {max (T) }}$ | 16.4\% | 1499 | $157$ | 2804 |
| F 2009 | 26.0\% | 1499 | 249 | 4446 |

Sound of Jura

|  | Harvest rate | Survey Index (adjusted) | Implied fishery |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Retained number | Landings (tonnes) |
| $\mathrm{F}_{\text {msy }}$ | 14.5\% | 251 | 23 | 515 |
| $\mathrm{F}_{2009}$ | 2.0\% | 251 | 3 | 71 |
| $\mathrm{F}_{0.1}(\mathrm{M})$ | 8.7\% | 251 | 14 | 309 |
| $\mathrm{F}_{0.1(\mathrm{~T})}$ | 9.7\% | 251 | 16 | 344 |
| $\mathrm{F}_{35 \% \mathrm{SpR}(\mathrm{M})}$ | 10.7\% | 251 | 17 | 380 |
| $\mathrm{F}_{\text {max }}(\mathrm{M})$ | 13.6\% | 251 | 22 | 483 |
| $\mathrm{F}_{35 \% \text { SpR(T) }}$ | 14.5\% | 251 | 23 | 515 |
| $\mathrm{F}_{\text {max ( }}(\mathrm{T})$ | 16.4\% | 251 | 26 | 582 |

$\mathrm{F}_{0.1(\mathrm{M}, \mathrm{T})}$ : Harvest ratio equivalent to fishing at a level associated with $10 \%$ of the slope at the origin on the male or combined sex YPR curve.
 bined SPR equal to $35 \%$ of the unfished level.
 combined YPR.

A discussion of $\mathrm{F}_{\mathrm{msy}}$ reference points for Nephrops is provided in Section 2.2.

### 3.7.9 Biological reference points

Precautionary approach biological reference points have not been determined for Nephrops stocks.

### 3.7.10 Uncertainties in the assessment and forecast

The length and sex composition of the landings data is considered to be well sampled. Discard sampling has been conducted on a quarterly basis for Scottish Nephrops trawlers in the Firth of Clyde subarea fishery since 1990, and is considered to represent the fishery adequately. Sampling in the Sound of Jura is sparser.

There are concerns over the accuracy of historical landings and effort data and because of this the final assessment adopted is independent of official statistics.

Underwater TV surveys have been conducted for this stock every year since 1995. The number of valid stations in the survey has remained relatively stable throughout the time period. Confidence intervals around the abundance estimates are stable throughout the series and relatively low compared with other FUs in VIa. There is a gap of 18 months between the survey and the start of the year for which the assessment is used to set management levels. It is assumed that the stock is in equilibrium during this period (i.e. recruitment and growth balance mortality) although this is rarely the case. The effect of this assumption on realised harvest rates has not been investigated.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. A three year average (2005-2007) of discard rate (adjusted to account for some survival of discarded animals) has been used in the calculation of catch options. Discard rates have fluctuated over the time-series but have been stable in the last two years. Mean weight has also fluctuated somewhat over the time-series. These uncertainties are not taken into account in the forecast.

The cumulative bias estimates for FU13 Clyde and Jura component is largely based on expert opinion (See Annex). The precision of these bias corrections cannot yet be characterised.

The survey should be considered as a minimum estimate. The overall area of the ground is estimated by contoured sediment data. New VMS data linked to landings suggests the area covered by the current UWTV is slightly smaller than the area covered by fishing activity and especially in the sea lochs. Figure 3.7.7 illustrates differences between the British Geological Survey based sediment approach to estimating area and the activity of $>15 \mathrm{~m}$ trawlers; inclusion of smaller vessels would likely further modify this. Work is in progress to refine the area estimate.

The landings forecast for 2011 (based on a transition value for the Firth of Clyde subarea and $\mathrm{F}_{\text {msy }}$ for the Sound of Jura subarea is almost 5806 tonnes. This is an increase on the reported landings in 2009 but below the peak values of 2007 and 2008.

### 3.7.11 State of stock

The perception of the state of the stock in the Firth of Clyde has not changed substantially since the assessment in 2008. The evidence from the TV survey suggests that the population is stable and the $15 \%$ decrease observed in 2009 is within the confidence limits for the past two years. The calculated harvest ratio in 2009 (dead removals/TV abundance) is above the values associated with high long-term yield and low risk depletion.

### 3.7.12 Management considerations

The WG, ACFM and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level
could confer controls to ensure effort and catch were in line with resources available. In this FU the two subareas imply that additional controls may be required to ensure that the landings taken in each subarea are in line with the landings advice. There is a need to reduce discards in this FU.

Creel fishing takes place in part of this area although the relative scale of the fishery is smaller than in the Minches. Overall effort in terms of creel numbers is not known and measures to control numbers are not in place. There is a need to ensure that the combined effort from all forms of fishing is taken into account when managing this stock.

There is a bycatch of other species in the area of the Firth of Clyde and STECF estimates that discards of whiting and haddock are high in VIa generally. It is important that efforts are made to ensure that unwanted bycatch is kept to a minimum in this fishery. Current efforts to reduce discards and unwanted bycatches of cod under the Scottish Conservation credits scheme and west coast emergency measures, include the implementation of larger meshed square meshed panels ( 120 mm ). A seasonal closure (early spring) in the southwest part of the Firth of Clyde is in place to protect spawning cod although Nephrops vessels are derogated to fish in those parts where mud sediments are distributed.

The implementation of buyers and sellers legislation in the UK in 2006 is improving the reliability of fishery statistics but the transition period is accompanied in some cases by large changes in landings which produce significant changes in the lpue and cpue series that cannot be completely attributed to changes in stock. Until a sufficient time-series of reliable data has built up, use of fishery catch and effort data in the assessment process should be avoided.

### 3.7.13 Other Nephrops populations within Division VIa

Nephrops fisheries also take place outside the Functional Units in Subdivision VIa, although they represent a low proportion of the reported landings (Table 3.5.3). Over the time-series, average landings have been just over 250 t and in recent 10 years, just over 300 t . An allowance for this activity is required in the final landings advice for 2011. The main areas of activity are the Stanton Bank (to the west of the South Minch, Figure 3.6.6) and areas of suitable sediment along the shelf edge and slope to the west of the Hebrides.

### 3.7.14 Stanton Bank

Underwater TV surveys were not conducted in Stanton Bank in 2008.

### 3.7.15 Shelf-edge west of Scotland

Marine Scotland Science has taken the opportunity of using the Scotia deep-water surveys conducted in 2000, 2002 and 2004 to conduct preliminary underwater TV work on the Nephrops populations along the shelf-edge. These TV runs are carried out during the night (when the vessel is not required for fishing). It is hoped that this can continue as an annual survey.

To date, successful survey runs have been conducted to a depth of 635 m , observing Nephrops burrows at a range of locations along the shelf-edge and slope. Observed densities have been very low (average $0.04 \mathrm{~m}^{-2}$ ) compared to shelf stocks on the west coast and in the North Sea (typically $0.2-0.9 \mathrm{~m}^{-2}$ ), although the animals on the shelfedge are considerably larger than those found on the shelf. Forecasts of landings based on TV surveys were not attempted for this area.

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Table 3.7.1. Nephrops, Clyde (FU13), Nominal Landings of Nephrops, 1981-2009, as officially reported.

| Year | UK Scotland |  |  |  | Other UK | Total ** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nephrops trawl | Other trawl | Creel | Sub-total |  |  |
| 1981 | 2498 | 404 | 66 | 2968 | 0 | 2968 |
| 1982 | 2373 | 171 | 79 | 2623 | 0 | 2623 |
| 1983 | 3890 | 120 | 53 | 4063 | 14 | 4077 |
| 1984 | 3069 | 154 | 77 | 3300 | 10 | 3310 |
| 1985 | 3921 | 293 | 64 | 4278 | 7 | 4285 |
| 1986 | 4074 | 175 | 79 | 4328 | 13 | 4341 |
| 1987 | 2859 | 80 | 65 | 3004 | 3 | 3007 |
| 1988 | 3507 | 108 | 43 | 3658 | 7 | 3665 |
| 1989 | 2577 | 184 | 35 | 2796 | 16 | 2812 |
| 1990 | 2732 | 122 | 24 | 2878 | 34 | 2912 |
| 1991 | 2845 | 145 | 25 | 3015 | 23 | 3038 |
| 1992 | 2532 | 246 | 10 | 2788 | 17 | 2805 |
| 1993 | 3199 | 110 | 5 | 3314 | 28 | 3342 |
| 1994 | 2503 | 49 | 28 | 2580 | 49 | 2629 |
| 1995 | 3767 | 132 | 26 | 3925 | 64 | 3989 |
| 1996 | 3880 | 111 | 27 | 4018 | 42 | 4060 |
| 1997 | 3486 | 44 | 25 | 3555 | 63 | 3618 |
| 1998 | 4539 | 81 | 40 | 4660 | 183 | 4843 |
| 1999 | 3475 | 29 | 38 | 3542 | 210 | 3752 |
| 2000 | 3143 | 63 | 76 | 3282 | 137 | 3419 |
| 2001 | 2889 | 67 | 94 | 3050 | 132 | 3182 |
| 2002 | 3074 | 53 | 105 | 3232 | 151 | 3383 |
| 2003 | 2954 | 20 | 117 | 3091 | 80 | 3171 |
| 2004 | 2659 | 18 | 90 | 2767 | 258 | 3025 |
| 2005 | 3166 | 14 | 95 | 3275 | 148 | 3423 |
| 2006 | 4446 | 0 | 0 | 4534 | 244 | 4778 |
| 2007 | 6129 | 0 | 0 | 6129 | 366 | 6495 |
| 2008 | 5291 | 29 | 182 | 5502 | 416 | 5918 |
| 2009* | 4277 | 26 | 190 | 4493 | 283 | 4776 |

* provisional ** Total also includes Rep. of Ireland.

Table 3.7.2. Nephrops, Clyde (FU13), Nominal Landings of Nephrops, in each of the subareas (Firth of Clyde and Sound of Jura 1981-2009, as officially reported.

|  | UK |  |  |
| :---: | :---: | :---: | :---: |
| Year | Firth of <br> Clyde | Sound <br> of Jura | All <br> sub-areas |
| 1981 |  |  | 2968 |
| 1982 |  |  | 2623 |
| 1983 |  |  | 4077 |
| 1984 |  |  | 3310 |
| 1985 |  |  | 4285 |
| 1986 |  |  | 3341 |
| 1987 |  | 3007 |  |
| 1988 |  | 3665 |  |
| 1989 |  | 2812 |  |
| 1990 |  | 2912 |  |
| 1991 |  | 3038 |  |
| 1992 |  |  | 2805 |
| 1993 | 2766 |  | 3342 |
| 1994 | 2094 | 535 | 2629 |
| 1995 | 3690 | 299 | 3989 |
| 1996 | 3673 | 387 | 4060 |
| 1997 | 3132 | 486 | 3618 |
| 1998 | 4372 | 471 | 4843 |
| 1999 | 3424 | 328 | 3752 |
| 2000 | 3230 | 189 | 3419 |
| 2001 | 2980 | 202 | 3182 |
| 2002 | 3349 | 34 | 3383 |
| 2003 | 3153 | 18 | 3171 |
| 2004 | 2975 | 50 | 3025 |
| 2005 | 3387 | 36 | 3423 |
| 2006 | 4717 | 61 | 4778 |
| 2007 | 6397 | 98 | 6495 |
| 2008 | 5840 | 78 | 5918 |
| $2009 *$ | 4684 | 92 | 4776 |
| * provisional | na = not available |  |  |
|  |  |  |  |

Table 3.7.3. Nephrops, Clyde (FU13): Firth of Clyde subarea. Mean sizes (CL mm) above and below 35 mm of male and female Nephrops in Scottish trawl catches and landings, 1981-2009.

| Year | Catches |  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | < 35 mm CL |  | $<35 \mathrm{~mm} \mathrm{CL}$ |  | > 35 mm CL |  |
|  | Males | Females | Males | Females | Males | Females |
| 1981 | 28.4 | 27.3 | 30.2 | 29.3 | 40.3 | 39.3 |
| 1982 | 28.2 | 26.4 | 29.9 | 29.0 | 39.9 | 40.1 |
| 1983 | 27.9 | 26.7 | 29.3 | 28.5 | 40.8 | 39.5 |
| 1984 | 27.0 | 25.9 | 28.0 | 26.8 | 40.9 | 39.6 |
| 1985 | 27.1 | 26.1 | 28.1 | 27.2 | 39.8 | 39.3 |
| 1986 | 27.1 | 26.0 | 27.9 | 27.1 | 40.5 | 39.0 |
| 1987 | 28.5 | 26.5 | 29.6 | 28.3 | 39.4 | 40.0 |
| 1988 | 28.1 | 27.0 | 30.6 | 29.5 | 41.2 | 40.1 |
| 1989 | 26.9 | 26.9 | 30.2 | 30.0 | 41.6 | 39.8 |
| 1990 | 27.4 | 26.2 | 30.4 | 29.5 | 40.1 | 39.8 |
| 1991 | 28.6 | 27.1 | 29.2 | 28.2 | 39.3 | 40.3 |
| 1992 | 29.6 | 28.8 | 30.1 | 29.2 | 39.9 | 41.1 |
| 1993 | 29.6 | 29.7 | 31.4 | 30.9 | 40.4 | 39.9 |
| 1994 | 26.4 | 27.0 | 29.4 | 29.4 | 40.8 | 39.2 |
| 1995 | 27.2 | 25.8 | 28.7 | 27.6 | 40.3 | 39.8 |
| 1996 | 28.8 | 28.0 | 30.0 | 29.1 | 38.6 | 40.4 |
| 1997 | 27.9 | 26.9 | 30.0 | 29.2 | 40.0 | 40.3 |
| 1998 | 25.9 | 25.2 | 28.4 | 27.9 | 38.9 | 39.1 |
| 1999 | 26.5 | 25.3 | 28.5 | 27.3 | 39.0 | 39.5 |
| 2000 | 28.3 | 27.7 | 29.3 | 28.6 | 38.7 | 39.1 |
| 2001 | 27.4 | 26.8 | 29.5 | 28.7 | 39.0 | 39.6 |
| 2002 | 27.5 | 25.6 | 28.4 | 26.4 | 39.0 | 39.4 |
| 2003 | 27.2 | 25.9 | 29.1 | 27.9 | 39.2 | 38.6 |
| 2004 | 27.1 | 26.5 | 28.4 | 27.6 | 39.2 | 39.5 |
| 2005 | 28.0 | 26.7 | 29.2 | 27.9 | 38.7 | 38.1 |
| 2006 | 28.7 | 27.1 | 29.0 | 27.3 | 40.0 | 38.7 |
| 2007 | 27.0 | 26.7 | 29.1 | 29.2 | 39.1 | 38.6 |
| 2008 | 27.2 | 25.2 | 28.6 | 26.6 | 39.1 | 38.2 |
| 2009 | 26.9 | 25.3 | 29.3 | 26.4 | 39.4 | 39.0 |

[^7]Table 3.7.4. Nephrops, Clyde (FU13): Firth of Clyde subarea. Results by stratum of the 2007-2009 TV surveys. Note that stratification was based on a series of sediment strata (M - Mud, SM Sandy mud, MS - Muddy sand).

| $\begin{aligned} & E \\ & \overrightarrow{3} \\ & \stackrel{0}{4} \\ & \stackrel{4}{n} \end{aligned}$ | $\begin{aligned} & \tilde{N} \\ & \stackrel{N}{\Sigma} \\ & \frac{1}{\Sigma} \\ & \hline \end{aligned}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 TV survey |  |  |  |  |  |  |  |
| M | 717 | 12 | 0.65 | 0.05 | 464 | 2344 | 0.159 |
| SM | 699 | 11 | 0.57 | 0.15 | 401 | 6553 | 0.445 |
| MS | 665 | 17 | 0.76 | 0.22 | 505 | 5812 | 0.395 |
| Total | 2081 | 40 |  |  | 1371* | 14709 | 1 |
| 2008 TV survey |  |  |  |  |  |  |  |
| M | 717 | 15 | 0.88 | 0.21 | 629 | 7345 | 0.173 |
| SM | 699 | 11 | 0.90 | 0.55 | 628 | 24502 | 0.575 |
| MS | 665 | 12 | 1.28 | 0.29 | 848 | 10732 | 0.252 |
| Total | 2081 | 38 |  |  | 2105 | 42579 | 1 |
| 2009 TV survey |  |  |  |  |  |  |  |
| M | 717 | 16 | 0.741 | 0.049 | 531 | 1583 | 0.102 |
| SM | 699 | 11 | 0.705 | 0.178 | 469 | 7150 | 0.459 |
| MS | 665 | 12 | 1.122 | 0.168 | 784 | 6842 | 0.439 |
| Total | 2081 | 39 |  |  | 1784 | 15575 | 1 |

*Note: abundance estimates for these years based on figures prior to the 2009 revision of the dat series. Differences between these figures and the revised figures shown on Table 3.7.5 are small.

Table 3.7.5. Nephrops, Clyde (FU13): Firth of Clyde subarea. Results of the 1995-2009 TV surveys.

|  | Stations | Mean density | Abundance | 95\% <br> confidence interval |
| :---: | :---: | :---: | :---: | :---: |
| Year |  | burrows/m ${ }^{2}$ | millions | millions |
| 1995 | 29 | 0.33 | 689 | 210 |
| 1996 | 38 | 0.54 | 1113 | 288 |
| 1997 | 31 | 0.68 | 1426 | 312 |
| 1998 | 38 | 0.720 | 1502 | 254 |
| 1999 | 39 | 0.532 | 1107 | 344 |
| 2000 | 40 | 0.807 | 1679 | 293 |
| 2001 | 39 | 0.850 | 1768 | 319 |
| 2002 | 36 | 0.899 | 1870 | 343 |
| 2003 | 37 | 1.039 | 2162 | 347 |
| 2004 | 32 | 1.127 | 2344 | 437 |
| 2005 | 44 | 1.121 | 2331 | 342 |
| 2006 | 43 | 1.050 | 2203 | 306 |
| 2007 | 40 | 0.705 | 1467 | 260 |
| 2008 | 38 | 1.012 | 2105 | 346 |
| 2009 | 39 | 0.86 | 1784 | 250 |

Table 3.7.6. Nephrops, Clyde (FU13): Sound of Jura subarea. Results by stratum of the 2009 TV surveys. Note that stratification was based on a series of sediment strata.

|  | 道 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 TV survey |  |  |  |  |  |  |  |
| M | 90 | 2 | 0.62 | 0.02 | 56 | 66 | 0.040 |
| SM | 142 | 5 | 1.18 | 0.28 | 168 | 1127 | 0.681 |
| MS | 150 | 5 | 0.50 | 0.10 | 75 | 463 | 0.279 |
| Total | 382 | 12 |  |  | 299 | 1656 | 1 |

Table 3.7.7. Nephrops, Clyde (FU13): Sound of Jura subarea. Results of the 1995-2009 TV surveys.


Table 3.7.8. Nephrops, Clyde (FU13): Firth of Clyde subarea. Adjusted TV survey abundance, landings, discard rate (proportion by number) and estimated harvest rate.

|  | Adjusted <br> survey <br> (millions) | Landings <br> (tonnes) | Discard rate (\%) | Harvest ratio* |
| :---: | :---: | :---: | :---: | :---: |
| 1999 | 930 | 3752 | 0.30 | 0.31 |
| 2000 | 1411 | 3419 | 0.22 | 0.15 |
| 2001 | 1486 | 3182 | 0.33 | 0.15 |
| 2002 | 1571 | 3383 | 0.19 | 0.16 |
| 2003 | 1817 | 3171 | 0.45 | 0.15 |
| 2004 | 1970 | 3025 | 0.52 | 0.15 |
| 2005 | 1959 | 3423 | 0.27 | 0.12 |
| 2006 | 1851 | 4778 | 0.18 | 0.16 |
| 2007 | 1233 | 6495 | 0.53 | 0.51 |
| 2008 | 1769 | 5918 | 0.37 | 0.29 |
| 2009 | 1499 | 4776 | 0.39 | 0.26 |

*harvest rates previous to 2006 are unreliable.

Table 3.7.9. Nephrops, Clyde (FU13): Sound of Jura subarea. Adjusted TV survey abundance, landings, discard rate (proportion by number) and estimated harvest rate.

|  | Adjusted survey <br> (millions) | Landings <br> (tonnes) | Discard rate (\%) | Harvest ratio* |
| :--- | :--- | :--- | :--- | :--- |
| 1999 | No Survey | 328 |  | No Survey |
| 2000 | No Survey | 189 |  | No Survey |
| 2001 | 13 | 202 |  |  |
| 2002 | 9 | 34 | 0.4 | No Survey |
| 2003 | 12 | 18 | 0.4 | 0.01 |
| 2004 | No Survey | 50 | 0.4 | 0.02 |
| 2005 | 303 | 36 | 0.4 | 0.03 |
| 2006 | 430 | 61 | 0.4 | No Survey |
| 2007 | 255 | 98 | 0.4 | 0.02 |
| 2008 | No Survey | 78 |  |  |
| 2009 | 251 | 92 |  |  |

*harvest rates previous to 2006 are unreliable.

## Landings - International



Figure 3.7.1. Nephrops, Clyde (FU13): Long-term landings, and mean sizes (Firth of Clyde subarea only).


Figure 3.7.2. Nephrops, Clyde (FU13), Firth of Clyde subarea, Landings, effort and lpues by quarter and sex from Scottish Nephrops trawlers.


Figure 3.7.3. Nephrops, Clyde (FU13), Catch length frequency distribution and mean sizes (red line) for Nephrops in the Firth of Clyde, 1979-2009.


Figure 3.7.4. Nephrops, Clyde (FU13), TV survey station distribution and relative density (burrows $/ \mathrm{m}^{2}$ ) for Firth of Clyde and Sound of Jura subareas, 2004-2009. Sound of Jura located to the east. Shaded green and brown areas represent areas of suitable sediment for Nephrops. Bubbles scaled the same. Red crosses represent zero observations.


Figure 3.7.5. Nephrops, Clyde (FU13): Firth of Clyde subarea. Time-series of revised TV survey abundance estimates (not adjusted for bias), with $95 \%$ confidence intervals.


Figure 3.7.6. Nephrops, Clyde (FU13): Sound of Jura subarea, Time-series of TV survey abundance estimates with $95 \%$ confidence intervals.


Figure 3.7.7. Nephrops, Clyde (FU13), Comparison of area of Nephrops ground defined by BGS sediment distribution (upper plot) and by distribution of VMS pings (shown in red) recorded from Nephrops trawlers $>15 \mathrm{~m}$ length (lower panel).

### 4.1 Rockall Area overview

There is no overview section.

### 4.2 Cod in Division VIb

Officially reported nominal landings are shown in Table 4.2.1 and Figure 4.2.1. Lpue results from the Irish otter trawl fleet are also presented in Figure 4.2.2. Figure 4.2.2 shows a large decline in lpue between 1995 and 2003 followed by relatively stable values at a level much lower than at the start of the time-series. No analytical assessment of this stock has been carried out.

Table 4.2.1. Cod in Division VIb (Rockall). Official catch statistics (nominal landings).

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 18 | - | 1 | - | 31 | 5 | - | - | - | 1 | - | - |
| France | 9 | 17 | 5 | 7 | 2 | - | - | - | - | - | - | - |
| Germany | - | 3 | - | - | 3 | - | - | 126 | 2 | - | - | - |
| Ireland | - | - | - | - | - | - | 400 | 236 | 235 | 472 | 280 | 477 |
| Norway | 373 | 202 | 95 | 130 | 195 | 148 | 119 | 312 | 199 | 199 | 120 | 92 |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - |
| Spain | 241 | 1200 | 1219 | 808 | 1345 | - | 64 | 70 | - | - | - | 2 |
| UK (E. \& W. \& N.I.) | 161 | 114 | 93 | 69 | 56 | 131 | 8 | 23 | 26 | 103 | 25 | 90 |
| UK (Scotland) | 221 | 437 | 187 | 284 | 254 | 265 | 758 | 829 | 714 | 322 | 236 | 370 |
| Total | 1,023 | 1,973 | 1,600 | 1,298 | 1,886 | 549 | 1,349 | 1,596 | 1,176 | 1,097 | 661 | 1,031 |
| Country | 19961 | 19971 | 1998 | 1999 | 2000 | 20012 | 2002 | 200 |  | 2004 | 42005 | 2006 |
| Faroe Islands | - | - | - | - | n/a | n/a | n/a |  |  |  |  |  |
| France | - | - | - | - | + | +* | 1 |  |  |  | 0.08 |  |
| Germany | 10 | 22 | 3 | 11 | 1 | - | - |  |  |  |  |  |
| Ireland | 436 | 153 | 227 | 148 | 119 | 40 | 18 |  | 11 | 7 | 12 | 22.7 |
| Norway | 91 | 55* | 51* | 85* | 152* | 89 | 28 |  | 25 | 23 | 7 | 7 |
| Portugal | - | 5 | - | - | - | - | - |  |  |  |  |  |
| Russia | - | - | - | - | 7 | 26 | - |  |  |  |  |  |
| Spain | 5 | 1 | 6 | 4 | 3 | 1 |  |  | 6 |  |  |  |
| UK (E. \& W. \& N.I.) | 23 | 20 | 32 | 22 | 4 | 2 | 2 |  | 3 |  |  |  |
| UK (Scotland) | 210 | 706 | 341 | 389 | 286 | 176 | 67 |  | 57 | 45 | 43 |  |
| UK |  |  |  |  |  |  |  |  |  |  |  | 28.7 |
| Total | 775 | 962 | 660 | 659 | 572 | 334 | 115 |  | 102 | 75 | 62 | 58.4 |


| Country | 2007 | 2008 | 2009* |
| :--- | :---: | :---: | :---: |
| Faroe Islands | - |  |  |
| France | - |  |  |
| Germany | - |  |  |
| Ireland | 24 | 40.7 | 20.4 |
| Norway | 12 | 11 | 25 |
| Portugal | - |  |  |
| Russia | - |  | 1 |
| Spain | - |  |  |
| UK (E. \& W. \& |  |  |  |
| N.I.) |  |  |  |
| UK (Scotland) | 26 |  |  |
| UK |  | 41.3 | 47.8 |
| Total | 62 | 93.0 | 94.2 |

[^8]

Figure 4.2.1. Cod in Division VIb. Total official nominal landings (all nations combined), 19842009.

Otter Trawl


Figure 4.2.2. Cod in Division VIb. Lpue from Irish Otter trawl fleet, 1995-2009.

### 4.3 Haddock in Division VIb (Rockall)

Type of assessment in 2010: Update assessment
The assessment of the haddock stock in Division VIb is based on catch-at-age and one survey index (Scottish Groundfish Survey) and conducted using the XSA method. Discarding occurs in part of the fishery. Discards have been estimated and used in the assessment. In 2005, WGNSDS, on the recommendation of RGNSDS, adopted a new assessment approach, which allows modelling of the total catch (including discards) of the Irish, Scottish and Russian fleets (for details see Stock Annex). The same approach has been used in the annual assessment since 2005. The current assessment is an update of the last year assessment.

## ICES advice applicable to 2009

The ICES advice for 2009 in terms of single-stock exploitation boundaries was as follows:

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects
"Fishing mortality around $F_{0.1}(0.21)$ can be considered as a candidate target reference point consistent with taking high long-term yields and achieving a low risk of depleting the productive potential ( $<5 \%$ ). The present fishing mortality (0.25) is above the candidate reference point."

Exploitation boundaries in relation to precautionary limits
"Fishing mortality should be less than $F_{p a}$, corresponding to total catches less than $9740 t$ in 2009. Landings should be less than 6470 t in 2009."

Conclusion on exploitation boundaries
"In the present situation with a stock that is well above $B_{p a}$ and fishing mortality below $F_{p a}$ there is little gain to the long-term yield by increasing fishing mortality above current levels. ICES therefore recommends to limit catches to 6490 t in 2009 and landings to $4330 t$."

## ICES advice applicable to 2010

The ICES advice for 2010 in terms of single-stock exploitation boundaries was as follows:

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects
"Fishing mortality around $F_{0.1}(0.18)$ can be considered as a candidate target reference point consistent with taking high long-term yields and achieving a low risk of depleting the productive potential ( $<5 \%$ ). The present fishing mortality (0.23) is above the candidate reference point and below $F_{p a .}$."

Exploitation boundaries in relation to precautionary limits
"Fishing mortality should be less than $F_{p a}$, corresponding to total catches less than 7090 tin 2010. Assuming that current discarding practices will be continued, landings should be less than 5480 t in 2010."

Considering the option below ICES advises that there is little gain on the long-term yield by increasing fishing mortality above current levels. ICES therefore recommends limiting catches and landings in 2010 to 4280 t and 3330 t , respectively.

### 4.3.1 General

## Stock description and management units

The haddock stock at Rockall is an entirely separate stock from that on the continental shelf of the British Isles. Since 2004, the EU TAC for haddock in VIb has been included with Divisions XII and XIV. For details of the earlier management units see Stock Annex.

## Management applicable to 2009 and 2010

The EU TAC for VIb, XII and XIV was set at 5879 t in 2009 (a $15 \%$ reduction compared to TAC for 2008).

The TAC for 2010 was set at 4997 t (a 15\% reduction compared to TAC for 2009) and is shown below:

| Species: | Haddock <br> Melanogrammus aeglefinus | Zone: | EU and international waters VIb, XII and XIV <br> (HAD/6B1214) |
| :--- | :--- | :--- | :--- |
| Belgium | 11 |  |  |
| Germany | 13 |  |  |
| France | 551 |  |  |
| Ireland | 393 |  |  |
| United Kingdom | 4029 | Analytical TAC |  |
| EU | 4997 |  |  |
| TAC | 4997 |  |  |

The ICES advice, agreed TAC for EU waters, and WG estimates of landings during 2002-2010 are summarised below. All values are in tonnes.

${ }^{\text {a }}$ TAC was set for Divisions VIa and VIb (plus Vb1, XII and XIV) combined with restrictions on quantity that can be taken in Vb and VIa. The quantity shown here is the total area TAC minus the maximum amount which is allowed to be taken from Vb and VIa.
${ }^{\text {b }}$ In 2004, the EU TAC for Division VI was split and the VIb TAC for haddock was included with XII and XIV. This value is the TAC for VIb, XII and XIV.
c Total catch, including landings and discards.
${ }^{d}$ Only landings.
The minimum landing size of haddock taken by EU vessels at Rockall is 30 cm . There is no minimum landing size for haddock taken by non-EU vessels in international waters.

In order to protect the pre-recruit stock, the International Waters component of the statistical rectangle 42D5 has been closed for fishing since 2001 and its EU component, since 2002 (see Stock Annex). The protected area (the whole rectangle) is referred to as Rockall Haddock Box. In order to protect cold-water corals, three further areas (North West Rockall, Logachev Mounds and West Rockall Mounds) were closed since January 2007 (see Stock Annex). A new area to protect cold-water corals (Empress of British Banks) was established by the NEAFC in 2007.

## Fishery in 2009

Nominal landings for 2009 and previous years as reported to ICES are given in Table 4.3.1.

## Russian fishery in 2009

In 2009, the fishery took place in September for only 7 days. 1 or 2 trawlers were engaged in the fishery (Table 4.3.2). Haddock accounted for $80-90 \%$ of the catch.

The vessels operated in the international waters at depths between 200 and 400 m . The total haddock catch was 55 t (Table 4.3.2), 53 t being taken by trawls and 2 t by longlines. This was a sharp decline compared to 1669 t taken in 2008.

## Scottish fishery in 2009

The number of Scottish vessels fishing for haddock and the number of trips made to Rockall declined substantially from 2000 onwards (WD6 to WGNSDS 2004). The declining trend was reversed in 2007. The number of vessels in increased from 22 in 2007 to 28 in 2008, and 37 in 2009.

The officially reported effort (in hours fishing) has varied over the past few years. However, these estimates are unreliable as reporting hours fishing is not mandatory for Scottish vessels and are not reported here for 2008 and 2009. Also, it is not known to what extent any variation in effort reflects changes in targeting haddock. Total Scottish demersal landings in VIb in 2009 are estimated to be 4585 t , of which 2951 t were haddock. The latter was an increase by $66 \%$ compared to the haddock catch in 2008 (Tables 4.3.1, 4.3.3). Other important target species included anglerfish (Lophius spp.), saithe, ling and megrim.

The UK landings and effort data included only Scottish vessels in 2009.

## Irish fishery in 2009

Landings totalling 352 t were reported from Irish otter trawlers in 2009 (over a twofold decrease from 721 t in 2008; Table 4.3.1). Most landings and effort were reported for Quarter 2 (Table 4.3.4).

## Norwegian fishery in 2008

The Norwegian demersal fleet fishing on the Rockall Bank consisted mainly of longliners and targeted mainly ling and tusk. Haddock constituted the bycatch in this fishery. All catch of haddock was taken in Quarters 3 and 4. In 2009, Norwegian landings of haddock amounted to 71 t which was a two-fold increase compared to 2008, and was within the catch range for the periods 2001-2005 and 2007-2009 (32-84 t).

### 4.3.2 Data

## Landings

Nominal landings as reported to ICES are given in Table 4.3.1, along with Working Group estimates of total estimated landings. Reported international landings of Rockall haddock in 1991-2005 varied between 4000 and 6000 t , except for 2001-2002, when they decreased down to about 2300-3000 t. In 2006, they were also low at 2760 t , but increased to 3348 t in 2007 and 4221 t in 2008. In 2009, international landings decreased to 3237 t .

Revisions to official catch statistics for previous years are also shown in Table 4.3.1.
Anecdotal evidence suggests that misreporting of haddock from Rockall has occurred historically (which may have led to discrepancies in assessment), but a quantitative estimation of the degree of misreporting is not possible.

Age composition and mean weight-at-age of Scottish and Irish landings were obtained from port sampling.

Age composition and mean weight-at-age of Russian landings were obtained by observers onboard commercial fishing vessels. In 2002, there was no sampling of the Russian catch and therefore the length composition for that year had to be estimated (for estimation details, see Stock Annex). In 2009, the Russian catch was down to 55 t and there was no sampling. The age composition in the Russian catch in 2009 was assumed to be the same as in the Scottish catches including discards.

Observer data from commercial vessels are also available for Norwegian landings for 2006-2009.

## Discards

Discarding by EC fleets is significant and therefore the assessment of the stock is done based on the total catch (landings+ discards). On Russian vessels, the whole catch of haddock is kept onboard and therefore, total catch is equivalent to landings.

Haddock discards onboard Scottish and Irish vessels were in some years determined directly, while in other years, indirect estimates of discards were done (for details of the estimation of discards see Stock Annex).

The analysis of the discard data collected by Scottish scientists in 1999 and 2001 indicated that only a relatively small proportion of fish taken aboard is landed (Figure 4.3.1). The direct estimates from the Scottish trawlers in 1985, 1999 and 2001 showed a higher proportion of discards of small haddock: from 12 to $75 \%$ by weight (Table 4.3 .5 ) and up to $80-90 \%$ of catch abundance. Discard trips in 1995, 1997, 1998, 2000 and 2001 showed that discarding by Irish fishing vessels is variable with a mean rate of $30 \%$ (Table 4.3.6).

Discard data were also obtained by Irish scientists from discard trips in 2007-2009. They showed that 52,87 and $63 \%$ of the catch in numbers, respectively, was discarded. The range of discarded sizes was 19-43 cm (mean 30 cm ) (Table 4.3.7). It should be noted that these estimates are based on very few trips (1, 2 and 3 for 2007, 2008 and 2009 respectively) and should therefore be treated with caution.

The proportion of fish discarded from Scottish and Irish catches at different sizes may be determined and modelled using a logistic curve. Calculations where the discard curve was applied agree well with the results of size composition measurements from Scottish vessels in 1999 and 2001 and from the combined 1995-2002 Irish discard trips (see Stock Annex).

Russian vessels retain all haddock and therefore there is no need to calculate discards (see Stock Annex).

There are some Scottish discards data for 2009, but their quality is very poor. Only six fish were measured, at age 4 and length $28-33 \mathrm{~cm}$.

## Biological

There was no change in biological parameters compared to the 2009 assessment (see Stock Annex).

## Surveys

There is only one abundance index available for VPA assessment of this stock from the Scottish survey (Figure 4.3.2). The survey is conducted in about 40 standard trawl stations. However, the survey area varied along with the number of stations in different years and survey covers only part of the currently known distribution area of haddock (see Stock Annex).

The distribution of sampling stations has slightly varied over time (Figure 4.3.2). The stations located in the southwest were not sampled every year and area what was covered by survey considerably differed in same years. Survey data were standardized for exploratory run in 2009. The stations which were located in the southwest were excluded from calculation. VPA was run with the old and new standardized indices (Tables 4.3.8, 4.3.9).

The Russian trawl-acoustic survey conducted in 2005 provided information on the stock size and biomass of the haddock stock, both in the EU zone and in international
waters. The acoustic survey yielded a biomass estimate of 60000 t and an abundance estimate of 225.9 million (for the details see Stock Annex). No such survey has been conducted in subsequent years.

## Commercial cpue

Commercial cpue series are available for Scottish trawlers, light trawlers, seiners, Irish otter trawlers and Russian trawlers fishing in Division VIb. The effort data for these five fleets are shown in Figure 4.3.3 and Table 4.3.10. Commercial cpue series for the different fleets are shown in Figure 4.3.4.

In 2005-2009, the Russian effort in bottom fishery (in hours and number of vessels/days) decreased due to economic reasons (Figure 4.3.4). Haddock catches varied accordingly with the changes in fishing effort. In 2006-2007, cpue in the Russian haddock fishery (mainly with trawlers of tonnage class 10) increased compared to previous years. In 2008-2009, it slightly decreased (with trawlers of class 8 and 9 only). The dynamics of catch per unit of effort for vessels agrees of tonnage class 10 agreed well with year-to-year variations in total biomass of haddock (Figure 4.3.5).

The effort data from the Scottish fleets are known to be unreliable due to changes in the practices of effort recording and non-mandatory effort reporting (see the Report of WGNSSK 2000, CM 2001/ACFM:07, for further details). It is unknown what proportion of Scottish and Irish effort was applied directly to the haddock fishery. The apparent effort increase may just be the result of more exact reporting of effort due to VMS, but another suggestion is that it arises from restrictive 'days at sea' in other areas (VIa and IV). Working at Rockall keeps 'days at sea' elsewhere intact (the years in question do correspond to the introduction of the days at sea legislation) and it is possible that vessels are either working extra days in VIb or they are simply reporting extra days from VIb. Despite the uncertainty about the fishing effort, the lpue for the Scottish fleet increased considerably in 2007 and 2008 compared to previous years (Figure 4.3.4).

The Irish otter trawl effort series indicated low values between 2002 and 2005 with the lowest value in 2004. In 2006-2008, the effort increased considerably, but declined in 2009 (Figure 4.3.3). The lpue showed an increase in 2007-2009 (Figure 4.3.4).

The WG decided that the commercial cpue and lpue data, which do not include discards and have not been corrected for changes in fishing power despite known changes in vessel size, engine power, fish-finding technology and net design, were unsuitable for catch-at-age tuning.

## Other relevant data

The Irish Fisheries Board (BIM) and the Marine Institute recently conducted a collaborative series of surveys to assess the length structure of haddock at various locations on the Rockall Bank and tested the selectivity of a number of codend configurations, which are typically used by the Irish fleets.

The selectivity of gears with different mesh sizes was also investigated at Rockall by Russian scientists.

### 4.3.3 Historical stock development

Model used:
The assessment is based on catch-at-age data and one survey index (Scottish Groundfish Survey) and conducted using the XSA method.

Software used:
The same software was used as in the last year's assessment (XSA from Lowestoft suite of VPA programs).

Model Options chosen:
Settings for the final XSA assessment did not change compared to the previous assessment (see Stock Annex) and were as follows:

Assessment model: XSA
Tuning indices: one survey index (SCOGFS)
Time-series weights: none
Catchability dependent for ages $<4$
Regression type: C
Q plateau: 5
Shrinkage stand. error: 1.0
Shrinkage age-year: 4 years, 3 ages
Minimum stand. error: 0.3
Plus group: 7+
Fbar: 2-5
Input data types and characteristics:
There were no changes in data types and characteristics compared to the previous assessment:

Year range: 1991-2009
Age range: 1-7+
For tuning data the following year and age ranges were used:
Year range: 1991-2009
Age range: 1-6

## Data screening

Figures 4.3.6 and 4.3.7 and Table 4.3.11 show landings, discards and total catch by number and weight. Landings, discards and total catch-at-age by number are shown in Tables 4.3.12-4.3.14.

Mean weights-at-age in total catch, landings, discards and stock are shown in Tables 4.3.15-4.3.18. The mean weights-at-age in the stock are assumed to be the same as the catch weights. The temporal dynamics of haddock mean weights-at-age in the total catch (including discards) and in the stock are shown in Figure 4.3.8. Mean weights-at-age in total catch were higher in 2008-2009 compared to 2007. This increase was observed in the Scottish landings and in the Russian catches in 2008 also.

The landings of haddock aged 1 were not large and it was hard to consider the catch of fish in this age group. The results from Scottish and Irish investigations showed that the abundance in discards exceeded that of landings. Discarded fish are, primarily, haddock aged 1-2 (see Tables 4.3.1 and 4.3.2 in Stock Annex). Figures of log catch by age show that these values are much less variable when discards are included
(Figures 4.3.9-4.3.14). Data on catches, landings and discards-at-age are given in Tables 4.3.12-4.3.14.

The Scottish trawl survey was the only survey index available to the Working Group. Plots of $\log$ cpue by age, year and year class are shown in Figures 4.3.15-4.3.17.

A SURBA 3.0 run was carried out to analyse the survey data. Previous working groups have concluded that the first three years of the survey should not be used in assessments and that age 0 data were a poor indicator of year-class strength. Here, the runs were actually conducted using the survey data from 1991 onwards to be consistent with the period over which the catch-at-age assessment could be run (the settings: lambda $=1.0$, reference age $=3$ ). A summary of the results are shown in Figure 4.3.18. SSB shows a declining trend from 1995, an increase in 2003-2004 and a general decrease in the subsequent years. The estimates of the temporal component of F are very noisy, but indicate a steep decline since 2000. Retrospective analysis showed consistent estimation of SSB and F (2-5) (Figure 4.3.18a).

Comparative scatter plots of log index-at-age are shown in Figure 4.3.18b. The survey shows relatively good internal consistency in tracking year-class strength through time.

## Final Update Assessment

## Exploratory runs

Two survey indices were used for XSA runs: old and new standardized indices. Both had rather small residuals in runs with data to 2008 (Figures 4.3.19-4.3.20 and Tables 4.3.19-4.3.20). However, tuning had not converged for the run with old original indices. In 2008 during the Scottish survey the age samples were collected by the Russian scientists. Indices for 2008 were calculated also with use of this age key also and tuning was converged in run with this indexes. However, the Russian age key is absent for 2009 and were used original 2009 indices. Standardization of indices on square covered by survey has allowed reducing log residuals for last year's especially.

However, including 2009 data negatively affected the quality of the assessment.
Tuning converged for a run with new standardized indices. However, log residuals for age 5 were high (Figure 4.3.21). There is a good correlation for other age classes. Big residuals in this assessment are the result of the strong year class 2005 and very poor year classes 2006-2009 (Table 4.3.21). The results of this assessment are presented in Table 4.3.22.

Tuning did not converge for a run with old original survey indices and log residuals for age 5 were high but less compare to new indices (Table 4.3.23, Figure 4.3.22, ). There is a good correlation for all age classes (Table 4.3.23).

For comparison, Figure 4.3 .23 shows SSB, recruitment at age 1 and mean F (2-5) estimates in the present assessment and assessments in which standardized survey indices were used. These estimates are rather consistent for the two index-series.

## Final run

In the final run, old original indices were used. The diagnostics file of the final XSA run is given in Table 4.3.23. Adjusted survey cpue against XSA population estimates are shown in Figures 4.3.24-4.3.25. The analysis of residuals and retrospective analysis (Figures 4.3.21-4.3.22, 4.3.26) shows that applying the chosen parameters for XSA (as done in 2005-2009 assessments) improves the residual patterns compared to other
exploratory settings. However, there are still same trends apparent in the log catchability residuals. The results of the retrospective analysis conducted by the Working Group in 2002 and 2003 indicated that using shrinkage values of more than 0.5 improved the retrospective curves and showed convergence. In this year's analysis, only 18 years data were available for the retrospective analysis, but a good year-to-year consistency was obtained. Dynamics of fishing mortality-at-age are presented in Figure 4.3.27. The final XSA results are given in Tables 4.3.24-4.3.26. The final XSA and SURBA results are compared in Figure 4.3.28. The SURBA estimates are more variable, but there is a good overall consistency between estimates by the two methods.

Summary plots from the final XSA assessment are shown in Figure 4.3.29.

## Comparison with previous assessments

XSA was conducted with the same basic assumptions and setup as last year's assessment. Perceptions of the stock have not changed. Figure 4.3 .30 shows, for comparison, SSB, recruitment at age 1 and mean $\mathrm{F}(2-5)$ estimates in the present assessment and assessments going back to 2001. The estimates from this year's assessment are reasonably consistent with the assessments carried out in previous years. Estimates of fishing mortality for 2009 have been revised upwards by $78 \%$, and SSB has been revised upwards by $1 \%$ (Figure 4.3.30).

## State of the stock

Based on this year's estimate of SSB and fishing mortality in 2009, the stock can be considered as having full reproductive capacity and that it is harvested sustainably. Spawning biomass has generally increased in recent years as a result of the 2001 and 2005 year classes. SSB has been above $B_{p a}$ since 2003. But SSB reduced in 2009. Fishing mortality was above $\mathrm{F}_{\mathrm{pa}}$ throughout most of the time-series but declined in 2005 and has remained below $\mathrm{F}_{\mathrm{pa}}$ since then.

## Statistical catch-at-age analysis (SCAA)

For Statistical catch-at-age analysis, StatCam model was used (J. Brodziak, 2005). VPA and SCAA used identical survey and catch data. For StatCam runs two scenarios were used. First scenario-non-parametric model, second-parametric model.

StatCam model shows good conformity between observed and predicted survey index and catch biomass. Log residuals were less 0.4 for total survey index (Figures 4.3.31-4.3.32).

StatCam summary plots are shown in Figure 4.3.33.
Both Statistical catch-at-age analysis and VPA results show a similar tendency for the SSB dynamics. However, the assessment of the stock size depends on the choice of the model. SSB and TSB plots from the XSA and SCAA assessment are compared in Figure 4.3.34.

### 4.3.4 Short-term projections

## Estimating year-class abundance

The abundance index for age 0 in the 2009 survey was low (Figure 4.3.35). VPA abundance for age 1 has been highly correlated with age 0 indices over most of the time-series (from 1993 onwards, Figure 4.3.36). The recruitment (age 1) in 2010 was
therefore estimated using RCT3 regression (Shepherd, 1997) relating survey indices to stock abundance.

For forecasting recruitment (age 1) in 2011 and thereafter, a geometric mean was used for 1991-2007.

The input data for the short-term forecast can be found in Table 4.3.27. Status quo fishing mortality is taken as a 3-year mean of the values over the period 2007-2009. Three year mean values were also used for stock weights and catch weights.

For forecasting discards and landings, the proportion of discards/landings-at-age in 1999-2009 was used, (Tables 4.3.11-4.3.14, Figure 4.3.37). The results obtained from the forecast (including discards) are given in Tables 4.3.27-4.3.29. The short-term forecast is also shown in Figure 4.3.38.

The sensitivity analysis of forecast is shown in Figures 4.3.39. There is a high probability of SSB in 2012 being below $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{F}_{\mathrm{sq}}$.

Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes are shown in Table 4.3.30.

### 4.3.5 Medium-term projection

Medium-term projections were conducted using the Marlab software. There appears to be little or no relationship between spawning biomass and recruitment levels at age 1 and no attempt to fit a stock-recruitment relationship to these data has been made. Particularly high discard rates result in very poor estimation of both the overall level and the inter-annual variability of recruitment. Significant year-to-year fluctuations of recruit abundance can be seen, and that the link between adult haddock biomass and abundance of survived fingerlings and yearlings is absent. In the years when biomass is at high levels, poor year classes are often observed. So in 2001, when the stock was low, one of the most abundant year classes appeared. Strong year classes appear on average once every $4-5$ years, although the available time-series is relatively short. SSB has been higher than $B_{p a}$ in recent years but recruitment for the last four years has been low which may be a consequence of rising temperature. With $\mathrm{F}_{\mathrm{sq}}=0.25$ for landings (total $\mathrm{F}_{\mathrm{sq}}=0.34$ ), there is a $30 \%$ probability of SSB falling below $\mathrm{B}_{\mathrm{pa}}$ in the long term (See Figures 4.3.40-4.3.42).

### 4.3.6 Biological reference points

Precautionary approach reference points
Biological reference points for this stock are given below:

| Blim: | $6,000 \mathrm{t}$ (lowest observed SSB) |
| :--- | :--- |
| $\mathrm{B}_{\mathrm{pa}}:$ | $9,000 \mathrm{t}\left(\mathrm{B}_{\text {loss }} \times 1.4\right)$ |
| $\mathrm{F}_{\mathrm{pa}}:$ | 0.4 (by analogy with other haddock stocks). |

Figure 4.3 .43 shows the stock in 2009 to be above $\mathrm{B}_{\mathrm{pa}}$ and below $\mathrm{F}_{\mathrm{pa}}$.

## Yield-per-recruit analysis

The stock-recruitment scatter plot is shown in Figure 4.3.44. Yield-per-recruit results, long-term yield and SSB (conditional on the current exploitation pattern) are shown
in Figure 4.3.45. Status quo F (0.34) is approximately $18 \%$ lower than $\mathrm{F}_{\max }(0.40)$ and twice as high as $\mathrm{F}_{0.1}$ (0.16).

## MSY evaluation

MSY estimates were evaluated using the srmsymc ADMB package. The number of stock and recruit pairs for this stock is fairly limited and these also show a relatively wide dynamic range. Yield and $\mathrm{F}_{\text {bar }}$ refer to total catch including landings and discards. Figures 4.3.46-4.3.50 show box plots of $\mathrm{F}_{\text {msy }}$ and $\mathrm{F}_{\text {crash }}$ as well as $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{Flim}_{\text {lim }}$. The deterministic fit lies outside the $5-95 \%$ percentiles for the Beverton and Holt S-R model which is therefore rejected. The hockey-stick breakpoint is poorly defined as us the Ricker curve. The Ricker model suggests that $\mathrm{F}_{\text {msy }}$ is above current F. The Ricker model assumes impaired recruitment at high stock levels and therefore arrives at higher $\mathrm{F}_{\mathrm{msy}}$ than the other models. This level of $\mathrm{F}_{\mathrm{msy}}$ overlaps with the confidence limits of $\mathrm{F}_{\text {crash. }}$. The yield plots for all three models suggest that MSY is relatively well defined although the absolute level of yield is highly uncertain and that current fishing mortality is close to $\mathrm{F}_{\text {msy }}$ estimate based on the underlying data.

Given the high CVs on all F parameters the WG concluded that the underlying data do not support the provision of absolute estimates of $\mathrm{F}_{\mathrm{msy}}$ but that current F was close that expected to deliver long-term equilibrium yield.

### 4.3.7 Management plans

There is a need for an internationally agreed management plan. This would require a management strategy evaluation to identify an appropriate $\mathrm{F}_{\mathrm{MSY}}$ target. Such a plan should involve extensive collaboration between stakeholders, scientists and management authorities in both the design and the monitoring of conservation measures. Management measures in the haddock fishery could be a combined application of TAC and limits of fishing efforts and should include effective control and enforcement measures. It would be beneficial to develop and introduce into fisheries practice measures aimed at minimising exploitation of juveniles.

In 2008-2009 the Russian Federation and the European Community have had consultations to develop a fisheries management plan. The report of the scientific working group was presented to the Delegations in 2009. It was recognised that the report contained all the relevant available data on the state of the stock and identified the issues, which would require continued cooperation between the Parties both at scientific and management levels.

In 2004, an ICES Expert Group met to deal with a request for advice from the EU and Russia concerning Rockall haddock management plans. They concluded that the lack of alternative assessment approaches precluded the identification of potential alternative limits to exploitation that may be useful to long-term management. In addressing this term of reference the Expert Group considered alternative approaches to management.

The 2004 Expert Group acknowledged that the Precautionary Approach requires that management be implemented in data poor situations. The Expert Group considered that the principles of the Precautionary Approach may have application to Rockall haddock provided the implementation considers the particular biology of the target species and the way it is exploited. For Rockall haddock the Expert Group considered that the fishing mortality should not be allowed to expand. Adoption of a TAC may actually allow increased fishing mortality if the stock is declining or there is significant unreported catch. Moreover, application of TACs implies that there is a simple
relationship between a recorded landing of a species and the effort exerted on that species. Such an assumption is unlikely to be true for Rockall haddock. Furthermore, there are ways of evading TACs including misreporting, high grading and discarding. In the case of Rockall haddock these may occur to a large extent due to the remote nature of the fishery and the processing of catches at sea by some fleets. The Expert Group concluded that effort regulation rather than TACs may be a better means of controlling fishing mortality on Rockall haddock in the long term but that TAC regulation could be used in the future if more objective and accurate biological and fishery information are routinely provided (ICES CM 2004/ACFM:33). In circumstances where population is dominated by small individuals and differences in length of older and younger age groups are not great, the effectiveness of using selective properties of trawl gear is very low. Comparison of the discard practices of the national fleets operating at Rockall indicate that an increase of minimum mesh size (as was the case in 1991) does not result in considerable reduction of the proportion of small individuals in catches, however catch rates are decreased. ACFM 2007 was unable to forecast discards and include them in TAC, and as a result, there were no recommendations on allowable landings. ACOM 2008 recommended applying TAC to landings only.

### 4.3.8 Uncertainties and bias in assessment and forecast

The WG considers that the long-term trends in the XSA assessment and survey biomass estimates/indices are probably indicative of the general stock trends. However, $F$ is considered to be poorly estimated due to the following sources of uncertainty in the current assessment:

1 ) The method of estimating discards from survey data, although considered appropriate, is likely to be the main source of error.
2 ) There are concerns over the accuracy of landings statistics from Rockall in earlier years.
3 ) Historically, there is poor agreement between survey and XSA estimates of population numbers during some periods. This may be related to potential inaccuracies in the landings statistics.
4 ) In 1999 the gear and tow duration were changed on the Scottish survey. There were no calibrations done to assess possible impacts on catchability for this survey.
5 ) The XSA assessment shows trends in catchability, even if reduced by weak shrinkage.
6 ) The XSA assessment diagnostics give quite large standard errors on survivors estimates (0.3-0.4) and there are often quite different values given by ScoGFS, F-shrinkage and P-shrinkage.

The WG considers that a longer series of more accurate landings, discards (for nonRussian fleets) and survey data will be necessary to overcome these deficiencies.

The survey covers only part of the currently known distribution area of haddock that raises uncertainty of an assessment.
There are concerns about the ability to forecast future catches and landings given substantial changes in national composition of the fleets operating at Rockall. A substantial change in TAC may lead to big changes in discarding practices. The Working Group previously presented forecast for total catch. However, with increased EU
catches with discards, this approach is no longer considered appropriate. The present forecast predicts future catches disaggregated into landings and discard components.

The WG makes the following reservations about the forecast:
1 ) The future fleet composition at Rockall is very uncertain
2 ) Discard proportion has varied considerably over time (Figure 4.3.37). However, no major changes in the pattern of discards-at-age have been observed since 1999, although this is based on few observations. Therefore, average proportions for 1999-2009 were used and it is assumed that these values will also apply for 2010-2012.

### 4.3.9 Recommendation for next Benchmark

The main conclusion of WGCSE is that and time-series of improved landings and discard data is needed before progress can be made towards the next benchmark assessment of this stock.

Because the survey covers only part of the currently known distribution area of haddock, it is necessary to use other available survey data for the assessment of this stock.

It is recommended to make the analysis of an opportunity of using of new models of an estimation including statistical catch-at-age analysis which will improve quality of assessment.

It would be beneficial to develop and introduce standardization methods for reading of age for haddock.

No timeframe for the next benchmark could be proposed at this stage.

### 4.3.10 Management considerations

Current fishing mortality is close to that which is expected to deliver long-term equilibrium yield. SSB in 2011 is higher than a $B_{p a}$ but the incoming recruitment for the last four years has been low and SSB is predicted to decline at current fishing mortality. Fishing at $\mathrm{F}_{\mathrm{pa}}$ in 2011 would result in a $26 \%$ reduction in SSB by 2012.

Fishing mortality levels have historically been high but have decreased since 2005. The fishing mortality has decreased for small individuals (age 1 and 2) since 2001. Survey-based indices of SSB indicate that the stock was at a historical low in 2002, but have increased since.

The forecast predicts future catches disaggregated into landing and discard components. The discard ratio is around $47 \%$ in 1991-2009 and $34 \%$ in the recent period (1999-2009). Some countries land the whole catch while others discard part of the catch. For countries which discard part of the catch the discard rate in the past was as high as $52-87 \%$ by numbers by results of discards trips. It would be beneficial to develop and introduce into fisheries practice measures aimed at preventing discards of haddock. Elaboration of such measures complies with recommendations under the UNGA Resolution 61/105 that urges states to take action to reduce or eliminate fish discards (UNGA Resolution 61/105, 2007, Chapter VIII, item 60).

In 2004-2009, the analytical methods of stock estimation were improved, the new data on biology and distribution were obtained, a trawl acoustic survey was carried out and the biomass of haddock from the Rockall Bank was estimated. The results from these investigations allow us to draw the following conclusions:

1) Due to the appearance of above-average year-classes in 2000-2001, the haddock stock has increased over the past few years. This is corroborated by Russian fishery statistics, biological research data, analytical calculations and Trawl Acoustic Survey in March 2005.
2 ) The 2005 year-class is also a strong one. It has grown to a catchable size and will enhance the fishable stock over the next few years.
3 ) It would be beneficial to conduct the ground fish/trawl-acoustic survey annually. An annual trawl survey covering the whole of the distributional area may improve the assessment of the stock status.
4 ) Discarding and the use of small-mesh gear have historically resulted in significant mortality of small haddock.
2) Regulation measures applied for haddock fishery encourage discards. Changes in the level of fishing mortality will not improve the situation as it will still be difficult to present forecasts both for discards and landings, and consequently for fishing mortality rates. Furthermore, there are ways of evading recommended fishing mortality including misreporting, high grading and discarding.
6 ) It would be beneficial to develop and introduce into fisheries practice measures aimed at preventing discards of undersized haddock.
3) General management issues aimed at maintaining a healthy stock of Rockall haddock, such as changes in landing size, changes in mesh size, use of square mesh and headline panels, licenses to fishing and closed areas, are currently being discussed through ongoing negotiations between EU and the Russian Federation.

### 4.3.11 References

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Table 4.3.1. Nominal catch (tonnes) of haddock in Division VIb, 1991-2009, as officially reported to ICES.

| Country | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 20091 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | - | - | - | - | n/a | n/a | - | - | - | - | 2 | 2 | 16 | - |
| France | $\ldots 2$ | $\ldots 2$ | $\ldots 2$ | $\ldots 2$ | ... 2 | - | - | - |  | 5 | 2 | - | 1 | - | - | - | - | - | - |
| Iceland | - | - | - | - | - | - | - | - | 167 | - | - | - | - | - | - | - | - | - | - |
| Ireland | 640 | 571 | 692 | 956 | 677 | 747 | 895 | 704 | 1,021 | 824 | 357 | 206 | 169 | 19 | 105 | 41 | 338 | 721 | 352 |
| Norway | 69 | 47 | 68 | 75 | 29 | 24 | 24 | 40 | 61 | 152 | 70 | 49 | 60 | 32 | 33 | 123 | 84 | 36 | 71 |
| Portugal | - | - | - | - | - | - | - | 4 | - | - | - | - | - | - | - | - | - | - | - |
| Russian Federation | - | - | - | - | - | - | - | - | 458 | 2,154 | 630 | 1,630 | 4,237 | 5,844 | 4,708 | 2,154 | 1,282 | 1669 | 55 |
| Spain | 187 | 51 | - | - | 28 | 1 | 22 | 21 | 25 | 47 | 51 | 7 | 19 | - | - | 5 | - | - | - |
| UK (E, W \& NI) | 165 | 74 | 308 | 169 | 318 | 293 | 165 | 561 | 288 | 36 | - | - | 56 | - | - | - | - | - | - |
| UK (Scotland) | 4,792 | 3,777 | 3,045 | 2,535 | 4,439 | 5,753 | 4,114 | 3,768 | 3,970 | 2,470 | 1,205 | 1,145 ${ }^{3}$ | 1,607 | $411^{3}$ | $332^{3}$ | $440^{3}$ | 1,643 ${ }^{3}$ | 1,779 ${ }^{3}$ | 2,951 ${ }^{3}$ |
| Total | 5,853 | 4,520 | 4,113 | 3,735 | 5,491 | 6,818 | 5,220 | 5,098 | 5,990 | 5,688 | 2,315 | 3,037 | 6,148 | 6,306 | 5,178 | 2,765 | 3,349 | 4,221 | 3,429 |
| Unallocated catch | -198 | 800 | 671 | 1,998 | -379 | -543 | -591 | -599 | -851 | -357 | -279 | 299 | 94 | 139 | 1 | 0 | 0 | 0 | -192 |
| WG estimate | 5,655 | 5,320 | 4,784 | 5,733 | 5,112 | 6,275 | 4,629 | 4,499 | 5,139 | $5,331^{4}$ | 2,036 ${ }^{4}$ | $3,336^{4}$ | $6.242^{4}$ | 6,445 | 5,179 | 2,765 | 3,349 | 4,221 | 3,237 |

${ }^{1}$ Preliminary.
${ }^{2}$ Included in Division VIa.
${ }^{3}$ Includes Scotland, England, Wales and NI landings
${ }^{4}$ includes the total Russian catch
n/a = not available.

Table 4.3.2. Details of Russian fleet operations in fishery for the haddock on the Rockall Bank (Division VIb) in 2009 (preliminary data).

| Month | Tonnage class | Number of <br> vessel/days | Catch of haddock, tonnes |
| :--- | :---: | :---: | :---: |
| September | 9 | 7 | Total |
| Total |  |  | 55 |

Table 4.3.3. Details of UK fleet operations in fishery for the haddock on the Rockall Bank (Division VIb) in 2009 (preliminary data).

| Month | Country | Gear type | Catch in tonnes |  |
| :--- | :--- | :--- | :---: | :---: |
|  |  |  | Total | Catch per vessel/day |
| February | Scotland | OTB | 147.6 | 18.5 |
| March | Scotland | OTB | 586.4 | 26.7 |
|  | Scotland | OTT | 7.3 | 7.3 |
| April | Scotland | OTB | 1031.3 | 30.3 |
|  | Scotland | OTT | 24.0 | 12.0 |
|  | Scotland | PTB | 142.5 | 17.8 |
| May | Scotland | OTB | 315.5 | 12.1 |
|  | Scotland | OTT | 40.1 | 8.0 |
|  | Scotland | PTB | 40.1 | 6.7 |
| June | Scotland | OTB | 149.6 | 8.8 |
|  | Scotland | OTT | 69.7 | 7.0 |
| July | Scotland | OTB | 60.2 | 5.5 |
|  | Scotland | OTT | 7.1 | 2.4 |
| August | Scotland | OTB | 52.4 | 7.5 |
| September | Scotland | OTB | 80.7 | 8.1 |
|  |  | OTT | 10.9 | 10.9 |
| October | Scotland | OTB | 70.3 | 7.0 |
| November | Scotland | OTB | 52.1 | 7.4 |
|  |  | OTT | 11.9 | 5.9 |
| December | Scotland | OTB | 53.7 | 17.9 |
| Total |  |  | 2951.3 |  |

OTB - bottom otter trawl, OTT - otter twin trawl, PTB - bottom pair trawl, SSC - Scottish seines.

Table 4.3.4. Details of Irish fleet operations in fishery for the haddock on the Rockall Bank (Otter Trawl. Division VIb) in 1995-2009 (preliminary data).

| Year | Landings (T) | Effort (HR*1000) | LPUE (KG/H) |
| :---: | :---: | :---: | :---: |
| 1995 | 839.99 | 9.14 | 91.88 |
| 1996 | 866.66 | 7.22 | 120.05 |
| 1997 | 830.99 | 7.17 | 115.91 |
| 1998 | 646.08 | 7.46 | 86.59 |
| 1999 | 973.64 | 8.68 | 112.17 |
| 2000 | 706.23 | 9.88 | 71.46 |
| 2001 | 300.47 | 7.24 | 41.48 |
| 2002 | 178.34 | 2.63 | 67.91 |
| 2003 | 155.93 | 4.56 | 34.23 |
| 2004 | 19.00 | 2.23 | 8.50 |
| 2005 | 103.54 | 3.84 | 26.93 |
| 2006 | 39.02 | 5.90 | 6.61 |
| 2007 | 340.84 | 6.59 | 51.73 |
| 2008 | 698.29 | 9.74 | 71.69 |
| 2009 | 349.44 | 4.35 | 80.26 |

Table 4.3.5. Details of Scottish discard trips in the Rockall area (Newton et al., 2003).

|  |  |  |  | \% (by weight) <br> haddock <br> landed of <br> catch | \% (by weight) <br> discarded of <br> haddock |  |
| :---: | :--- | :--- | :---: | :---: | :---: | :---: |
| Trip no. | Date | Gear | No. of <br> hauls | Hours <br> fished | May 85 | Heavy <br> Trawl |
| 1 | Jun 85 | Heavy <br> Trawl | 20 | 89.08 | 74 | 17.3 |
| 3 | Jun 99 | Heavy <br> Trawl | 21 | 110.83 | 41 | 18.17 |
| 4 | Apr 01 | Heavy <br> Trawl | 11 | 47.33 | 96 | 74.9 |
| 5 | Jun 01 | Heavy <br> Trawl | 35 | 163.58 | 58 | 12.4 |
| 6 | Aug 01 | Heavy <br> Trawl | 26 | 130.08 | 31 | 69.5 |

Table 4.3.6. Landings and Discards haddock estimates at Rockall from discard observer trips conducted aboard Irish vessels between 1995 and 2001, and from an observer trip aboard the MFV (February-March 2000). (ICES CM 2004/ACFM:33).

|  | FAT/ | FAT/ | FAT/ | FAT/ | FAT/ | FAT/ | FAT/ |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | KBG/ | KBG/ | KBG/ | KBG/ | KBG/ | KBG/ | KBG/ | Feb | Discard |
|  | $00 / 4$ | $01 / 12$ | $95 / 1$ | $95 / 2$ | $97 / 7$ | $97 / 8$ | $98 / 4$ | 2000 | rate |
| Landing | 3021 | 942 | 12727 | 6893 | 14258 | 25866 | 23805 | 4400 |  |
| Discards | 1864 | 926 | 1146 | 1893 | 6625 | 17926 | 3687 | 6200 |  |
| $\%$ |  |  |  |  |  |  |  |  | $27 \%$ |
| discarded | 38.16 | 49.57 | 8.26 | 21.54 | 31.72 | 40.90 | 13.40 | 58.49 |  |

Table 4.3.7. Discards and retained catch haddock (number per trip) by Irish discard trips in the Rockall area in 2007-2009.

| Year <br> Length (cm) | 2007 |  | 2008 |  | 2009 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Discards | Retained Catch | Discards | Retained Catch | Discards | Retained Catch |
| 19 | 1.3 |  |  |  |  |  |
| 22 | 1.6 |  | 14.8 |  |  |  |
| 23 | 4.6 |  | 66.2 |  |  |  |
| 24 | 7.3 |  | 183.8 |  |  |  |
| 25 | 22.7 |  | 576.9 |  | 15.6 |  |
| 26 | 54.2 |  | 1424.9 |  | 30.4 |  |
| 27 | 104.6 |  | 3024.6 |  | 25.2 |  |
| 28 | 256.9 |  | 6274.7 |  | 228.2 |  |
| 29 | 386.5 | 7.9 | 7193.3 |  | 180.6 |  |
| 30 | 533.4 | 17.6 | 7813.5 | 13.9 | 573.2 | 9.9 |
| 31 | 462.6 | 47.2 | 7573.7 | 40.6 | 1338.1 | 9.9 |
| 32 | 298.8 | 88.3 | 4639.0 | 77.8 | 1762.8 | 57.8 |
| 33 | 227.3 | 99.4 | 3664.7 | 126.8 | 2256.5 | 235.9 |
| 34 | 120.8 | 139.2 | 2391.8 | 277.4 | 1496.5 | 397.3 |
| 35 | 78.3 | 118.8 | 1590.1 | 503.6 | 656.6 | 614.8 |
| 36 | 27.4 | 187.0 | 871.7 | 580.5 | 423.5 | 567.1 |
| 37 | 26.1 | 139.8 | 280.3 | 640.9 | 66.9 | 526.8 |
| 38 | 24.3 | 142.7 | 78.3 | 581.9 | 57.4 | 421.4 |
| 39 | 3.4 | 162.5 | 206.6 | 443.0 | 23.1 | 346.9 |
| 40 | 8.7 | 119.4 | 37.5 | 535.6 |  | 281.4 |
| 41 | 1.3 | 133.8 | 5.2 | 310.7 |  | 197.9 |
| 42 | 4.6 | 133.1 | 5.2 | 334.7 |  | 155.7 |
| 43 | 3.2 | 109.3 |  | 333.5 |  | 195.1 |
| 44 |  | 118.6 |  | 291.1 |  | 201.7 |
| 45 |  | 97.9 |  | 253.6 |  | 149.9 |
| $>45 \mathrm{~cm}$ |  | 574.5 | 0.0 | 1791.2 | 0.0 | 1001.7 |
| Total | 2659.9 | 2436.9 | 47916.8 | 7136.8 | 9134.4 | 5371.3 |
| Discard rate, \% | 52.2 |  | 87.0 |  | 63.0 |  |

Table 4.3.8. Haddock in VIb. Old tuning data available from the Scottish groundfish survey conducted in September.

HADDOCK WGNSDS 2008 ROCKALL
101
SCOGFS (Numbers per 10 hours fishing at Rockall)
19912009
110.660 .75

06

| 1 | 14458 | 16398 | 4431 | 683 | 315 | 228 | 37 | 64 | 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 20336 | 44912 | 14631 | 3150 | 647 | 127 | 200 | 4 | 32 |
| 1 | 15220 | 37959 | 15689 | 3716 | 1104 | 183 | 38 | 73 | 21 |
| 1 | 23474 | 13287 | 11399 | 4314 | 969 | 203 | 30 | 12 | 4 |
| 1 | 16923 | 16971 | 6648 | 5993 | 1935 | 483 | 200 | 16 | 0 |
| 1 | 33578 | 19420 | 5903 | 1940 | 1317 | 325 | 69 | 6 | 1 |
| 1 | 28897 | 10693 | 2384 | 538 | 292 | 281 | 71 | 9 | 1 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 10178 | 9969 | 2410 | 708 | 279 | 172 | 90 | 64 | 32 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 31813 | 7455 | 521 | 284 | 154 | 39 | 14 | 12 | 14 |
| 1 | 11704 | 20925 | 2464 | 173 | 105 | 65 | 20 | 10 | 15 |
| 1 | 2526 | 10114 | 10927 | 1656 | 138 | 97 | 100 | 26 | 6 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1 | 24452 | 4082 | 920 | 1506 | 2107 | 231 | 33 | 13 | 7 |
| 1 | 3570 | 18715 | 2562 | 256 | 1402 | 1694 | 349 | 16 | 6 |
| 1 | 558 | 2671 | 6019 | 570 | 254 | 516 | 367 | 28 | 2 |
| 1 | 85 | 560 | 966 | 3813 | 182 | 41 | 282 | 249 | 49 |
| 1 | 132 | 139 | 323 | 488 | 1651 | 40 | 9 | 54 | 17 |

Table 4.3.9. Haddock in VIb. Standardized tuning data available from the Scottish groundfish survey conducted in September.

HADDOCK WGCSE 2010 ROCKALL
101
SCOGFS
19912009

| 1 | 1 | 0.66 | 0.75 |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 6 |  |  |  |  |  |  |  |  |
| 1 | 14838 | 16830 | 4548 | 701 | 323 | 234 | 38 | 66 | 2 |
| 1 | 10347 | 22748 | 7489 | 1614 | 331 | 65 | 103 | 2 | 16 |
| 1 | 16268 | 36664 | 15653 | 3867 | 1156 | 193 | 40 | 76 | 22 |
| 1 | 22921 | 12509 | 10893 | 4210 | 956 | 200 | 31 | 13 | 3 |
| 1 | 17650 | 16775 | 6011 | 5155 | 1699 | 430 | 176 | 14 | -1 |
| 1 | 33586 | 19424 | 5908 | 1945 | 1324 | 329 | 69 | 6 | 1 |
| 1 | 28910 | 10697 | 2395 | 544 | 299 | 295 | 76 | 11 | 1 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |
| 1 | 10138 | 8773 | 2372 | 706 | 265 | 169 | 94 | 60 | 21 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |
| 1 | 31808 | 7425 | 520 | 285 | 154 | 39 | 14 | 12 | 4 |
| 1 | 11703 | 20925 | 2463 | 172 | 105 | 65 | 20 | 10 | 15 |
| 1 | 2526 | 10114 | 10928 | 1656 | 138 | 97 | 101 | 26 | 6 |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |  |
| 1 | 24450 | 4078 | 920 | 1509 | 2105 | 233 | 35 | 12 | 8 |
| 1 | 1675 | 8890 | 1561 | 158 | 815 | 973 | 200 | 10 | 5 |
| 1 | 558 | 2671 | 6019 | 570 | 254 | 516 | 367 | 28 | -1 |
| 1 | 84 | 222 | 378 | 3401 | 1217 | 371 | 164 | 76 | 82 |
| 1 | 134 | 125 | 286 | 445 | 1546 | 38 | 8 | 54 | 17 |

Table 4.3.10. Details of Scottish and Irish effort (in hours) in 1985-2009 (preliminary data)

| Year | Scottish fleet |  |  | Irish fleet IROTB* |
| :---: | :---: | :---: | :---: | :---: |
|  | SCOTRL* | SCOLTR* | SCOSEI* |  |
| 1985 | 8421 | 3081 | 1677 |  |
| 1986 | 7465 | 4783 | 507 |  |
| 1987 | 8786 | 9737 | 402 |  |
| 1988 | 12450 | 5521 | 261 |  |
| 1989 | 10161 | 11946 | 1411 |  |
| 1990 | 3249 | 5335 | 4552 |  |
| 1991 | 2995 | 11464 | 6733 |  |
| 1992 | 2402 | 9623 | 3948 |  |
| 1993 | 1632 | 11540 | 1756 |  |
| 1994 | 2305 | 15543 | 399 |  |
| 1995 | 1789 | 13517 | 1383 | 9142 |
| 1996 | 1627 | 17324 | 952 | 7219 |
| 1997 | 563 | 16096 | 1061 | 7169 |
| 1998 | 1332 | 12263 | 456 | 7461 |
| 1999 | 11336 | 9424 | 456 | 8680 |
| 2000 | 12951 | 8586 | 80 | 9883 |
| 2001 | 7838 | 1037 | 42 | 7244 |
| 2002 | 8304 | 1100 | 0 | 2626 |
| 2003 | 15000 | 500 | 50 | 4618 |
| 2004 | 15200 | 300 | 50 | 2070 |
| 2005 | 7788 | 32 | 0 | 2693 |
| 2006 | 9990 | 231 | 0 | 5903 |
| 2007 | 4534 | 319 | 44 | 6589 |
| 2008 | 2497 | 1016 | 82 | 9740 |
| 2009 | NA | NA | NA | 4354 |

SCOTRL* - Scottish Heavy Trawl , SCOLTR* - Scottish Light Trawl , SCOSEI* - Scottish Seine, IROTB* - Irish bottom otter trawl.

Table 4.3.11. Haddock in VIb International landings, discards and total catch.

|  | Num (*1000) |  |  |  |  |  |  | Weight, tonnes |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Discards | Total Catch1 | Landings | Discards | Total Catch ${ }^{1}$ |  |  |  |  |  |  |
| 1991 | 12302 | 65832 | 78134 | 5656 | 13228 | 18884 |  |  |  |  |  |  |
| 1992 | 11418 | 55964 | 67383 | 5321 | 11871 | 17192 |  |  |  |  |  |  |
| 1993 | 8767 | 44656 | 53423 | 4781 | 9853 | 14634 |  |  |  |  |  |  |
| 1994 | 11400 | 46628 | 58028 | 5732 | 11023 | 16755 |  |  |  |  |  |  |
| 1995 | 11784 | 35467 | 47251 | 5587 | 9168 | 14756 |  |  |  |  |  |  |
| 1996 | 14066 | 41506 | 55572 | 7072 | 9356 | 16428 |  |  |  |  |  |  |
| 1997 | 9965 | 26980 | 36945 | 5167 | 5894 | 11061 |  |  |  |  |  |  |
| 1998 | 9034 | 47831 | 56865 | 4986 | 10862 | 15848 |  |  |  |  |  |  |
| 1999 | 12930 | 52881 | 65811 | 5356 | 11062 | 16418 |  |  |  |  |  |  |
| 2000 | 15999 | 26033 | 42031 | 5444 | 6609 | 12053 |  |  |  |  |  |  |
| 2001 | 5361 | 9222 | 14583 | 2123 | 1535 | 3658 |  |  |  |  |  |  |
| 2002 | 11167 | 21899 | 33066 | 3117 | 4152 | 7270 |  |  |  |  |  |  |
| 2003 | 24409 | 25087 | 49496 | 5969 | 5521 | 11490 |  |  |  |  |  |  |
| 2004 | 22705 | 3989 | 26694 | 6437 | 883 | 7321 |  |  |  |  |  |  |
| 2005 | 19505 | 1877 | 21382 | 5191 | 505 | 5696 |  |  |  |  |  |  |
| 2006 | 9605 | 1667 | 11272 | 2756 | 386 | 3142 |  |  |  |  |  |  |
| 2007 | 8936 | 12261 | 21197 | 3348 | 2242 | 5590 |  |  |  |  |  |  |
| 2008 | 10209 | 7603 | 17812 | 4221 | 2100 | 6320 |  |  |  |  |  |  |
| 2009 | 6709 | 4765 | 11474 | 3237 | 1557 | 4794 |  |  |  |  |  |  |

${ }^{1}$ Landings and discards.

Table 4.3.12. Haddock in VIb. International catch (landings and discards) numbers $\left({ }^{*} 10^{3}\right)$-at-age.


Table 4.3.13. Haddock in VIb. International landings numbers $\left({ }^{*} 10^{3}\right)$-at-age.
Run title : HADDOCK LANDISC 2007 ROCKALL
At 16/05/2010 13:00

Terminal Fs derived using XSA (With F shrinkage)

| Landings number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| AGE |  |  |  |  |  |  |  |  |
|  | 1 | 87 | 86 | 28 | 30 | 1 | 2 | 0 |
|  | 2 | 6807 | 3642 | 1919 | 1160 | 146 | 5149 | 319 |
|  | 3 | 3011 | 5624 | 4740 | 5299 | 5205 | 1861 | 2102 |
|  | 4 | 1344 | 964 | 1157 | 3665 | 4791 | 4149 | 2155 |
|  | 5 | 558 | 580 | 489 | 1040 | 1319 | 2347 | 3658 |
|  | 6 | 32 | 364 | 144 | 66 | 279 | 473 | 1540 |
|  | +gp | 464 | 160 | 290 | 141 | 43 | 85 | 192 |
| 0 | TOT, | 12302 | 11418 | 8767 | 11400 | 11784 | 14066 | 9965 |


| Landings number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 4 | 245 | 33 | 399 | 657 | 920 | 197 | 887 | 2344 | 31 | 17 | 5 |
|  | 2 | 392 | 2600 | 3445 | 941 | 2983 | 8103 | 1765 | 2835 | 768 | 1220 | 749 | 11 |
|  | 3 | 1815 | 2994 | 5081 | 1232 | 3998 | 11001 | 9502 | 6866 | 1290 | 2709 | 6191 | 244 |
|  | 4 | 1340 | 1972 | 3006 | 752 | 2111 | 1846 | 9119 | 7913 | 2356 | 1074 | 1164 | 5243 |
|  | 5 | 1898 | 1228 | 1295 | 988 | 809 | 1188 | 1364 | 725 | 2269 | 1539 | 479 | 460 |
|  | 6 | 2284 | 1600 | 1176 | 470 | 217 | 878 | 286 | 98 | 428 | 1623 | 761 | 261 |
|  | +gp | 1301 | 2291 | 1963 | 579 | 392 | 475 | 472 | 182 | 150 | 740 | 848 | 486 |
| 0 | TOT, | 9034 | 12930 | 15999 | 5361 | 11167 | 24409 | 22705 | 19505 | 9605 | 8936 | 10209 | 6709 |

Table 4.3.14. Haddock in VIb. International discards numbers $\left({ }^{*} 10^{3}\right)$-at-age.

Run title : HADDOCK LANDISC 2007 ROCKALL At 16/05/2010 13:00

Terminal Fs derived using XSA (With F shrinkage)

| Discards number at age (start of year) |  |  |  | Numbers*10**-3 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1991 | 1992 | 1993 | 1994 | 1995* | 1996 | 1997* |
| AGE |  |  |  |  |  |  |  |
| 1 | 21099 | 15998 | 11151 | 8140 | 2748 | 12094 | 9957 |
| 2 | 27040 | 21069 | 17456 | 19464 | 9685 | 13662 | 10216 |
| 3 | 12178 | 12961 | 10755 | 12570 | 16379 | 9051 | 3286 |
| 4 | 3998 | 4397 | 3781 | 4545 | 4965 | 5463 | 1944 |
| 5 | 1146 | 1181 | 1128 | 1409 | 1145 | 952 | 1344 |
| 6 | 313 | 312 | 317 | 410 | 508 | 278 | 218 |
| +gp | 58 | 46 | 69 | 91 | 36 | 7 | 15 |
| 0 TOT, | 65832 | 55964 | 44656 | 46628 | 35467 | 41506 | 26980 |


| Discards number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1998 | 1999* | 2000 | 2001* | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 14220 | 17037 | 8189 | 7268 | 12706 | 5655 | 735 | 174 | 536 | 1459 | 458 | 218 |
|  | 2 | 19415 | 19348 | 9136 | 1019 | 8136 | 15503 | 2346 | 888 | 707 | 8610 | 1458 | 696 |
|  | 3 | 8357 | 9209 | 5616 | 583 | 539 | 3558 | 781 | 554 | 336 | 896 | 5246 | 993 |
|  | 4 | 3423 | 3526 | 1912 | 266 | 334 | 217 | 93 | 210 | 58 | 429 | 128 | 2803 |
|  | 5 | 1842 | 2191 | 755 | 50 | 89 | 97 | 22 | 28 | 22 | 674 | 28 | 35 |
|  | 6 | 483 | 1084 | 322 | 15 | 43 | 48 | 10 | 11 | 8 | 193 | 203 | 2 |
|  | +gp | 91 | 485 | 103 | 21 | 51 | 8 | 2 | 11 | 1 | 0 | 82 | 18 |
| 0 | TOT, | 47831 | 52881 | 26033 | 9222 | 21899 | 25087 | 3989 | 1877 | 1667 | 12261 | 7603 | 4765 |

* data calculated using estimates from discard observer trips.

Table 4.3.15. Haddock in VIb. International catch (landings and discards) weights-at-age (kg).

|  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.679 |
| 1992 | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.844 |
| 1993 | 0.137 | 0.238 | 0.334 | 0.400 | 0.493 | 0.503 | 0.874 |
| 1994 | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.721 |
| 1995 | 0.118 | 0.222 | 0.309 | 0.401 | 0.501 | 0.460 | 0.843 |
| 1996 | 0.136 | 0.278 | 0.314 | 0.395 | 0.553 | 0.575 | 0.763 |
| 1997 | 0.136 | 0.240 | 0.322 | 0.382 | 0.512 | 0.634 | 0.944 |
| 1998 | 0.141 | 0.250 | 0.308 | 0.354 | 0.436 | 0.546 | 0.662 |
| 1999 | 0.138 | 0.208 | 0.272 | 0.334 | 0.379 | 0.483 | 0.618 |
| 2000 | 0.189 | 0.250 | 0.267 | 0.321 | 0.382 | 0.451 | 0.707 |
| 2001 | 0.133 | 0.257 | 0.320 | 0.416 | 0.432 | 0.521 | 0.713 |
| 2002 | 0.135 | 0.239 | 0.237 | 0.325 | 0.509 | 0.580 | 0.753 |
| 2003 | 0.153 | 0.203 | 0.256 | 0.350 | 0.384 | 0.424 | 0.753 |
| 2004 | 0.147 | 0.198 | 0.244 | 0.294 | 0.444 | 0.609 | 0.753 |
| 2005 | 0.114 | 0.197 | 0.235 | 0.311 | 0.459 | 0.600 | 0.806 |
| 2006 | 0.093 | 0.198 | 0.245 | 0.329 | 0.441 | 0.595 | 0.787 |
| 2007 | 0.114 | 0.186 | 0.266 | 0.296 | 0.387 | 0.497 | 0.569 |
| 2008 | 0.199 | 0.241 | 0.291 | 0.437 | 0.571 | 0.669 | 0.932 |
| 2009 | 0.248 | 0.288 | 0.339 | 0.391 | 0.668 | 0.513 | 1.005 |

Table 4.3.16. Haddock in VIb. International landings weights-at-age (kg).

|  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | 0.302 | 0.402 | 0.444 | 0.592 | 0.724 | 0.963 | 0.704 |
| 1992 | 0.136 | 0.366 | 0.455 | 0.658 | 0.612 | 0.759 | 0.954 |
| 1993 | 0.305 | 0.402 | 0.503 | 0.701 | 0.830 | 0.820 | 0.972 |
| 1994 | 0.314 | 0.356 | 0.452 | 0.558 | 0.638 | 1.224 | 0.890 |
| 1995 | 0.377 | 0.311 | 0.414 | 0.479 | 0.640 | 0.699 | 1.236 |
| 1996 | 0.327 | 0.436 | 0.501 | 0.487 | 0.627 | 0.709 | 0.783 |
| 1997 | - | 0.315 | 0.401 | 0.444 | 0.564 | 0.661 | 0.973 |
| 1998 | 0.256 | 0.344 | 0.494 | 0.517 | 0.542 | 0.591 | 0.678 |
| 1999 | 0.274 | 0.338 | 0.390 | 0.440 | 0.505 | 0.601 | 0.665 |
| 2000 | 0.272 | 0.404 | 0.379 | 0.407 | 0.473 | 0.513 | 0.740 |
| 2001 | 0.274 | 0.426 | 0.383 | 0.518 | 0.426 | 0.518 | 0.677 |
| 2002 | 0.240 | 0.422 | 0.416 | 0.541 | 0.565 | 0.649 | 0.818 |
| 2003 | 0.100 | 0.164 | 0.246 | 0.351 | 0.388 | 0.423 | 0.758 |
| 2004 | 0.142 | 0.172 | 0.241 | 0.293 | 0.446 | 0.617 | 0.754 |
| 2005 | 0.103 | 0.184 | 0.230 | 0.310 | 0.461 | 0.614 | 0.824 |
| 2006 | 0.084 | 0.167 | 0.223 | 0.327 | 0.440 | 0.598 | 0.789 |
| 2007 | 0.096 | 0.238 | 0.275 | 0.322 | 0.450 | 0.523 | 0.570 |
| 2008 | 0.125 | 0.197 | 0.302 | 0.444 | 0.583 | 0.752 | 0.984 |
| 2009 | 0.300 | 0.346 | 0.420 | 0.416 | 0.692 | 0.512 | 1.020 |

Table 4.3.17. Haddock in VIb. International discards weights-at-age (kg).

|  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | 0.142 | 0.199 | 0.253 | 0.306 | 0.345 | 0.358 | 0.478 |
| 1992 | 0.133 | 0.217 | 0.258 | 0.298 | 0.330 | 0.342 | 0.464 |
| 1993 | 0.137 | 0.220 | 0.260 | 0.307 | 0.346 | 0.359 | 0.462 |
| 1994 | 0.153 | 0.226 | 0.263 | 0.308 | 0.345 | 0.356 | 0.458 |
| 1995 | 0.118 | 0.220 | 0.276 | 0.325 | 0.341 | 0.329 | 0.379 |
| 1996 | 0.136 | 0.218 | 0.276 | 0.326 | 0.370 | 0.348 | 0.524 |
| 1997 | 0.136 | 0.238 | 0.272 | 0.312 | 0.372 | 0.442 | 0.568 |
| 1998 | 0.141 | 0.248 | 0.267 | 0.291 | 0.327 | 0.336 | 0.436 |
| 1999 | 0.139 | 0.212 | 0.255 | 0.288 | 0.313 | 0.318 | 0.410 |
| 2000 | 0.189 | 0.267 | 0.289 | 0.311 | 0.330 | 0.334 | 0.462 |
| 2001 | 0.135 | 0.247 | 0.294 | 0.344 | 0.412 | 0.440 | 0.495 |
| 2002 | 0.137 | 0.254 | 0.308 | 0.335 | 0.398 | 0.338 | 0.367 |
| 2003 | 0.161 | 0.223 | 0.287 | 0.342 | 0.337 | 0.440 | 0.510 |
| 2004 | 0.148 | 0.218 | 0.282 | 0.343 | 0.324 | 0.371 | 0.469 |
| 2005 | 0.171 | 0.240 | 0.298 | 0.357 | 0.387 | 0.473 | 0.506 |
| 2006 | 0.132 | 0.233 | 0.334 | 0.420 | 0.495 | 0.435 | 0.435 |
| 2007 | 0.115 | 0.179 | 0.239 | 0.232 | 0.244 | 0.280 | 0.406 |
| 2008 | 0.202 | 0.264 | 0.279 | 0.370 | 0.351 | 0.358 | 0.392 |
| 2009 | 0.247 | 0.287 | 0.319 | 0.343 | 0.360 | 0.662 | 0.593 |

Table 4.3.18. Haddock VIb. Stock weights-at-age (kg).

|  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | 0.142 | 0.240 | 0.291 | 0.378 | 0.469 | 0.414 | 0.679 |
| 1992 | 0.133 | 0.239 | 0.318 | 0.362 | 0.423 | 0.567 | 0.844 |
| 1993 | 0.137 | 0.238 | 0.334 | 0.400 | 0.493 | 0.503 | 0.874 |
| 1994 | 0.153 | 0.233 | 0.319 | 0.420 | 0.469 | 0.477 | 0.721 |
| 1995 | 0.118 | 0.222 | 0.309 | 0.401 | 0.501 | 0.460 | 0.843 |
| 1996 | 0.136 | 0.278 | 0.314 | 0.395 | 0.553 | 0.575 | 0.763 |
| 1997 | 0.136 | 0.240 | 0.322 | 0.382 | 0.512 | 0.634 | 0.944 |
| 1998 | 0.141 | 0.250 | 0.308 | 0.354 | 0.436 | 0.546 | 0.662 |
| 1999 | 0.138 | 0.208 | 0.272 | 0.334 | 0.379 | 0.483 | 0.618 |
| 2000 | 0.189 | 0.250 | 0.267 | 0.321 | 0.382 | 0.451 | 0.707 |
| 2001 | 0.133 | 0.257 | 0.320 | 0.416 | 0.432 | 0.521 | 0.713 |
| 2002 | 0.135 | 0.239 | 0.237 | 0.325 | 0.509 | 0.580 | 0.753 |
| 2003 | 0.153 | 0.203 | 0.256 | 0.350 | 0.384 | 0.424 | 0.753 |
| 2004 | 0.147 | 0.198 | 0.244 | 0.294 | 0.444 | 0.609 | 0.753 |
| 2005 | 0.114 | 0.197 | 0.235 | 0.311 | 0.459 | 0.600 | 0.806 |
| 2006 | 0.093 | 0.198 | 0.245 | 0.329 | 0.441 | 0.595 | 0.787 |
| 2007 | 0.114 | 0.186 | 0.266 | 0.296 | 0.387 | 0.497 | 0.569 |
| 2008 | 0.199 | 0.241 | 0.291 | 0.437 | 0.571 | 0.669 | 0.932 |
| 2009 | 0.248 | 0.288 | 0.339 | 0.391 | 0.668 | 0.513 | 1.005 |

Table 4.3.19. Regression statistics. Old survey indexes. Run to year 2008.
Regression statistics:
Ages with q dependent on year class strength

| Age | Slope |  | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Log |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |
| 1 | 0.61 | 3.239 | 5.19 | 0.84 | 15 | 0.28 | -1.48 |  |
| 2 | 0.68 | 2.07 | 4.77 | 0.77 | 15 | 0.37 | -2.04 |  |
| 3 | 0.57 | 4.136 | 5.79 | 0.88 | 15 | 0.25 | -2.51 |  |

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value | Intercept | RSquare | No Pts | Reg s.e |  |
| ---: | :--- | ---: | :--- | ---: | ---: | ---: | ---: | ---: | Mean Q

Table 4.3.20. Regression statistics. Standardized survey indexes. Run to year 2008.

Ages with q dependent on year class strength

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Log |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.53 | 3.671 | 6.02 | 0.83 | 15 | 0.29 | -1.64 |
| 2 | 0.54 | 4.036 | 6.09 | 0.86 | 15 | 0.25 | -2.21 |
| 3 | 0.49 | 5.4 | 6.47 | 0.89 | 15 | 0.2 | -2.61 |

Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |
| 4 | 0.6 | 2.926 | 5.26 | 0.8 | 15 | 0.27 | -2.56 |
| 5 | 0.98 | 0.1 | 2.72 | 0.57 | 15 | 0.51 | -2.57 |
| 6 | 1.04 | -0.536 | 2.43 | 0.94 | 15 | 0.22 | -2.62 |

Table 4.3.21. XSA diagnostics in assessment of Haddock in VIb. Exploratory runs with standardized survey index.

## Lowestoft VPA Version 3.1

14/05/2010 20:40
Extended Survivors Analysis
HADDOCK LANDISC 2009 ROCKALL
CPUE data from file had6b.tun

Catch data for 19 years. 1991 to 2009. Ages 1 to 7.


Time series weights :

Tapered time weighting not applied
Catchability analysis :
Catchability dependent on stock size for ages < 4
Regression type $=\mathrm{C}$
Minimum of 10 points used for regression
Survivor estimates shrunk to the population mean for ages < 4
Catchability independent of age for ages >= 5
Terminal population estimation :

Survivor estimates shrunk towards the mean F
of the final 4 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.000$

Minimum standard error for population
estimates derived from each fleet $=.300$
Prior weighting not applied
Tuning converged after 29 iterations
1

| Regression weights |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.387 | 0.118 | 0.15 | 0.173 | 0.075 | 0.047 | 0.043 | 0.081 | 0.051 | 0.031 |
| 2 | 0.851 | 0.148 | 0.251 | 0.429 | 0.156 | 0.476 | 0.085 | 0.202 | 0.166 | 0.1 |
| 3 | 1.022 | 0.27 | 0.598 | 0.609 | 0.336 | 0.464 | 0.394 | 0.306 | 0.383 | 0.132 |
| 4 | 1.233 | 0.232 | 0.714 | 0.607 | 1.044 | 0.486 | 0.267 | 0.787 | 0.171 | 0.512 |
| 5 | 1.231 | 0.986 | 0.33 | 1.103 | 1.156 | 0.203 | 0.243 | 0.42 | 0.68 | 0.091 |
| 6 | 1.272 | 1.204 | 0.724 | 0.676 | 0.838 | 0.235 | 0.173 | 0.309 | 0.326 | 0.958 |

Table 4.3.21 cont.

XSA population numbers (Thousands)

|  | AGE |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 |

2000 2.83E+04 2.43E+04 1.85E+04 7.67E+03 3.20E+03 2.30E+03 2001 7.61E+04 1.58E+04 8.48E+03 5.44E+03 1.83E+03 7.65E+02 $2002 \quad 1.06 \mathrm{E}+05 \quad 5.54 \mathrm{E}+04 \quad 1.11 \mathrm{E}+04 \quad 5.30 \mathrm{E}+03 \quad 3.53 \mathrm{E}+03 \quad 5.59 \mathrm{E}+02$ 2003 4.58E+04 7.48E+04 3.53E+04 5.01E+03 2.12E+03 2.08E+03 2004 1.43E+04 3.15E+04 3.98E+04 1.57E+04 2.24E+03 5.77E+02 $2005 \quad 2.57 \mathrm{E}+04 \quad 1.09 \mathrm{E}+04 \quad 2.21 \mathrm{E}+04 \quad 2.33 \mathrm{E}+04 \quad 4.53 \mathrm{E}+03 \quad 5.76 \mathrm{E}+02$ 2006 7.57E+04 2.01E+04 5.52E+03 1.14E+04 1.17E+04 3.03E+03 2007 2.11E+04 $5.94 \mathrm{E}+04 \quad 1.51 \mathrm{E}+04 \quad 3.05 \mathrm{E}+03 \quad 7.13 \mathrm{E}+03 \quad 7.54 \mathrm{E}+03$
2008 1.06E+04 1.59E+04 3.97E+04 9.10E+03 1.14E+03 3.83E+03
2009 8.19E+03 8.25E+03 1.11E+04 2.22E+04 6.28E+03 4.71E+02

Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+006.50 \mathrm{E}+036.11 \mathrm{E}+037.93 \mathrm{E}+031.09 \mathrm{E}+044.69 \mathrm{E}+03$
Taper weighted geometric mean of the VPA populations:
$4.68 \mathrm{E}+04 \quad 3.74 \mathrm{E}+04 \quad 2.29 \mathrm{E}+04 \quad 1.10 \mathrm{E}+04 \quad 4.45 \mathrm{E}+03 \quad 1.69 \mathrm{E}+03$
Standard error of the weighted Log(VPA populations) :

```
\(\begin{array}{llllll}0.8212 & 0.7205 & 0.619 & 0.5841 & 0.6538 & 0.8684\end{array}\)
1
```

Log catchability residuals.
Fleet: SCOGFS

| Age |  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | -0.26 | -0.12 | -0.01 | 0.01 | 0.25 | 0.39 | -0.1 | 99.99 | 0.31 |  |
|  | 2 | -0.34 | 0.07 | 0.42 | -0.03 | 0.19 | 0.35 | -0.16 | 99.99 | -0.14 |  |
|  | 3 | -0.3 | -0.1 | 0.39 | 0.2 | 0.06 | 0.04 | -0.45 | 99.99 | -0.08 |  |
|  | 4 | -0.17 | -0.08 | 0.51 | 0.47 | 0.67 | -0.05 | -1.13 | 99.99 | -0.37 |  |
|  | 5 | 0.25 | -0.46 | 0.71 | -0.33 | 0.88 | 0.09 | -0.64 | 99.99 | -0.29 |  |
|  | 6 | 0.07 | 0.12 | 0.03 | -0.06 | 0.14 | -0.1 | -0.34 | 99.99 | -0.11 |  |
| Age |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 99.99 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 2 | 99.99 | -0.37 | -0.14 | 0.32 | 99.99 | 0.36 | -0.3 | 0.34 | -0.31 | -0.37 |
|  | 3 | 99.99 | 0.11 | -0.5 | 0.24 | 99.99 | 0.6 | 0.15 | -0.02 | -0.5 | -0.05 |
|  | 4 | 99.99 | -0.81 | -0.82 | -0.12 | 99.99 | 0.23 | 0.24 | -0.04 | 0.09 | 0.05 |
|  | 5 | 99.99 | -0.38 | -0.99 | -0.57 | 99.99 | 0.53 | 0.15 | 0.66 | 0.7 | 0.29 |
|  | 6 | 99.99 | -0.38 | -0.05 | 0.22 | 99.99 | -0.05 | 0.45 | 0.44 | 2.13 | -2.27 |
|  |  |  |  |  |  |  |  | 0.14 | 0.18 | -0.03 | -0.15 |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 4 | 5 | 6 |
| :---: | ---: | ---: | ---: |
| Mean Log | -2.4536 | -2.63 | -2.63 |
| S.E(Log q | 0.5981 | 0.9524 | 0.2374 |

Regression statistics :
Ages with q dependent on year class strength

| Age | Slope |  | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Log q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |
|  | 1 | 0.54 | 4.972 | 5.96 | 0.89 | 16 | 0.3 | -1.76 |
| 2 | 0.64 | 3.267 | 5.21 | 0.86 | 16 | 0.33 | -2.23 |  |
|  | 3 | 0.59 | 4.466 | 5.6 | 0.9 | 16 | 0.23 | -2.59 |

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4 | 0.79 | 1.051 | 3.86 | 0.65 | 16 | 0.47 | -2.45 |
|  | 5 | 2.33 | -1.616 | -5.04 | 0.1 | 16 | 2.1 | -2.63 |
|  | 6 | 0.94 | 0.94 | 2.98 | 0.94 | 16 | 0.22 | -2.69 |

Table 4.3.21 cont.

Terminal year survivor and F summaries :
Age 1 Catchability dependent on age and year class strength
Year class $=2008$


| Survivors <br> at end of J | Int | Ext | N |  | Var | F |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 6502 | 0.31 | s.e |  |  | Ratio |  |

Age 2 Catchability dependent on age and year class strength
Year class $=2007$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N |  | Scaled <br> Weights | Estimatec <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SCOGFS | 5072 | 0.255 | 0.134 | 0.52 |  | 2 | 0.789 | 0.119 |
| P shrinki | 22901 | 0.62 |  |  |  |  | 0.152 | 0.028 |
| F shrinké | 2434 | 1 |  |  |  |  | 0.058 | 0.233 |
| Weighted prediction : |  |  |  |  |  |  |  |  |


| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 6113 | 0.23 | 0.36 |  | 4 | 1.571 |  |

Age 3 Catchability dependent on age and year class strength
Year class $=2006$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimatec <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SCOGFS | 8027 | 0.187 | 0.223 | 1.19 | 3 | 0.851 | 0.13 |
| P shrink | 10989 | 0.58 |  |  |  | 0.111 | 0.097 |
| F shrinke | 2350 | 1 |  |  |  | 0.038 | 0.389 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors at end of ) | s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | N | Var Ratio | F |  |  |
| 7934 | 0.18 | 0.2 | 5 | 1.111 | 0.132 |  |  |

Table 4.3.21. cont.

Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2005$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimatec <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SCOGFS | 10703 | 0.178 | 0.113 | 0.64 | 4 | 0.925 | 0.519 |
| F shrinke | 13516 | 1 |  |  |  | 0.075 | 0.431 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of ) | s.e | s.e |  | Ratio |  |  |  |
| 10892 | 0.18 | 0.1 | 5 | 0.551 | 0.512 |  |  |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=2004$

| Fleet |  | Int | Ext | Var | N | Scaled |  | Estimatec |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | s.e | s.e | Ratio |  |  | Weights | F |
| SCOGFS | 5109 | 0.174 | 0.294 | 1.69 |  | 5 | 0.95 | 0.084 |
| F shrinke | 942 | 1 |  |  |  |  | 0.05 | 0.389 |

Weighted prediction :


1
Age 6 Catchability constant w.r.t. time and age (fixed at the value for age) 5
Year class $=2003$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimatec <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SCOGFS | 109 | 0.233 | 0.32 | 1.37 | 5 | 0.85 | 1.16 |
| F shrinke | 852 | 1 |  |  |  | 0.15 | 0.246 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of $)$ | s.e | s.e |  | Ratio |  |  |  |
| 148 | 0.25 | 0.44 | 6 | 1.785 | 0.958 |  |  |

Table 4.3.22 Haddock in VIb. Exploratory runs with standardized survey indices. Summary tables.

Run title : HADDOCK LANDISC 2004 ROCKALL
At 14/05/2010 20:41

Terminal Fs derived using XSA (With F shrinkage)

| Table 8 Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.2395 | 0.1763 | 0.1046 | 0.1404 | 0.0506 | 0.2404 | 0.1663 | 0.2429 | 0.4975 |  |  |
|  | 2 | 0.6045 | 0.4869 | 0.3337 | 0.2858 | 0.2503 | 0.5689 | 0.3413 | 0.579 | 0.7302 |  |  |
|  | 3 | 0.8944 | 0.8137 | 0.6547 | 0.5909 | 0.5497 | 0.4872 | 0.3121 | 0.6533 | 0.8919 |  |  |
|  | 4 | 0.9229 | 0.9761 | 0.5237 | 0.9121 | 0.7705 | 0.5084 | 0.3397 | 0.5034 | 0.9381 |  |  |
|  | 5 | 0.5677 | 0.9437 | 0.9402 | 0.5394 | 0.7898 | 0.653 | 0.5468 | 0.5994 | 0.8525 |  |  |
|  | 6 | 0.5893 | 0.4622 | 0.6974 | 0.8236 | 0.3295 | 0.5937 | 0.9156 | 0.6766 | 1.2715 |  |  |
|  | +gp | 0.5893 | 0.4622 | 0.6974 | 0.8236 | 0.3295 | 0.5937 | 0.9156 | 0.6766 | 1.2715 |  |  |
| 0 | FBAR 2 | 0.7474 | 0.8051 | 0.6131 | 0.5821 | 0.5901 | 0.5544 | 0.385 | 0.5838 | 0.8531 |  |  |
|  | Table 8 | Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR **-** |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.3866 | 0.118 | 0.1499 | 0.1729 | 0.0748 | 0.0467 | 0.0429 | 0.0812 | 0.0508 | 0.0305 | 0.0542 |
|  | 2 | 0.8514 | 0.1479 | 0.2509 | 0.4292 | 0.1556 | 0.4764 | 0.0847 | 0.2019 | 0.166 | 0.0996 | 0.1558 |
|  | 3 | 1.0221 | 0.27 | 0.5984 | 0.609 | 0.3357 | 0.464 | 0.394 | 0.3065 | 0.3827 | 0.1319 | 0.2737 |
|  | 4 | 1.233 | 0.2316 | 0.7136 | 0.6071 | 1.0442 | 0.4861 | 0.2673 | 0.7873 | 0.1706 | 0.5119 | 0.49 |
|  | 5 | 1.2315 | 0.9864 | 0.3296 | 1.1033 | 1.1562 | 0.203 | 0.2428 | 0.4203 | 0.6797 | 0.0912 | 0.3971 |
|  | 6 | 1.2724 | 1.2042 | 0.7235 | 0.6763 | 0.8377 | 0.2347 | 0.1734 | 0.3094 | 0.3258 | 0.9585 | 0.5312 |
|  | +gp | 1.2724 | 1.2042 | 0.7235 | 0.6763 | 0.8377 | 0.2347 | 0.1734 | 0.3094 | 0.3258 | 0.9585 |  |
| 0 | FBAR 2 | 1.0845 | 0.409 | 0.4731 | 0.6871 | 0.6729 | 0.4074 | 0.2472 | 0.429 | 0.3498 | 0.2086 |  |

Table 4.3.22. cont. Run title : HADDOCK LANDISC 2010 ROCKALL At 14/05/2010 20:41

Terminal Fs derived using XSA (With F shrinkage)

|  | Table 10 <br> YEAR | Stock number at age (start of year) |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 109945 | 109958 | 124345 | 68946 | 61567 | 62555 | 71834 | 72897 | 48729 |  |  |  |  |
|  | 2 | 82461 | 70846 | 75473 | 91690 | 49056 | 47919 | 40270 | 49803 | 46813 |  |  |  |  |
|  | 3 | 28398 | 36887 | 35645 | 44261 | 56409 | 31269 | 22211 | 23438 | 22854 |  |  |  |  |
|  | 4 | 9795 | 9506 | 13385 | 15164 | 20070 | 26653 | 15728 | 13310 | 9985 |  |  |  |  |
|  | 5 | 4347 | 3186 | 2933 | 6491 | 4987 | 7604 | 13125 | 9168 | 6587 |  |  |  |  |
|  | 6 | 858 | 2017 | 1015 | 938 | 3099 | 1853 | 3240 | 6220 | 4122 |  |  |  |  |
|  | +gp | 1281 | 610 | 781 | 451 | 308 | 224 | 374 | 3087 | 4169 |  |  |  |  |
| 0 | TOTAI | 237084 | 233012 | 253576 | 227940 | 195495 | 178076 | 166783 | 177924 | 143259 |  |  |  |  |
|  | Table 10 | Stock number at age (start of year) |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | GMST 91-** | AMS |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 28342 | 76112 | 106080 | 45772 | 14290 | 25683 | 75748 | 21121 | 10600 | 8188 | 0 | 56592 | 66113 |
|  | 2 | 24258 | 15765 | 55378 | 74759 | 31525 | 10856 | 20068 | 59412 | 15944 | 8248 | 6502 | 42952 | 49785 |
|  | 3 | 18467 | 8477 | 11133 | 35279 | 39848 | 22090 | 5519 | 15095 | 39748 | 11057 | 6113 | 23140 | 26899 |
|  | 4 | 7669 | 5441 | 5298 | 5011 | 15710 | 23321 | 11372 | 3048 | 9097 | 22195 | 7934 | 10662 | 12380 |
|  | 5 | 3200 | 1830 | 3533 | 2125 | 2235 | 4527 | 11743 | 7126 | 1135 | 6279 | 10892 | 4728 | 5573 |
|  | 6 | 2299 | 765 | 559 | 2081 | 577 | 576 | 3026 | 7542 | 3832 | 471 | 4693 | 1739 | 2399 |
|  | +gp | 3102 | 928 | 940 | 1072 | 909 | 1014 | 1045 | 3055 | 3672 | 887 | 426 |  |  |
| 0 | TOTAI | 87338 | 109317 | 182921 | 166098 | 105095 | 88068 | 128521 | 116399 | 84029 | 57325 | 36561 |  |  |

Table 4.3.22. cont.


Table 4.3.23. XSA diagnostics in assessment of Haddock in VIb. Final run with old survey indices.

Lowestoft VPA Version 3.1
4/06/2010 9:33

Extended Survivors Analysis
HADDOCK LANDISC 2004 ROCKALL
CPUE data from file had6b.tun

Catch data for 19 years. 1991 to 2009. Ages 1 to 7 .

| Fleet |  |  | First |  | Last |  | Alpha | Beta |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | year | ear | age |  | age |  |  |  |
| SCOGFS | 1991 | 2009 |  | 0 |  | 6 | 0.66 |  |

Time series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability dependent on stock size for ages < 4

Regression type $=C$
Minimum of 10 points used for regression
Survivor estimates shrunk to the population mean for ages < 4

Catchability independent of age for ages $>=5$

Terminal population estimation :

Survivor estimates shrunk towards the mean F of the final 4 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.000$

Minimum standard error for population
estimates derived from each fleet $=.300$
Prior weighting not applied

Tuning had not converged after 180 iterations

Total absolute residual between iterations 179 and $180=.01111$

| Final year F values |  |  |  | 5 | 6 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1 | 2 | 3 | 4 | 5 | 0.214 |
| Iteration * | 0.038 | 0.0975 | 0.1319 | 0.423 | 0.9048 |  |
| Iteration ** | 0.0377 | 0.097 | 0.1313 | 0.4226 | 0.2086 | 0.901 |

Regression weights

| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.385 | 0.111 | 0.144 | 0.152 | 0.075 | 0.067 | 0.039 | 0.081 | 0.05 | 0.038 |
| 2 | 0.844 | 0.147 | 0.232 | 0.409 | 0.134 | 0.474 | 0.125 | 0.184 | 0.165 | 0.097 |
| 3 | 1.01 | 0.266 | 0.593 | 0.541 | 0.313 | 0.38 | 0.391 | 0.504 | 0.337 | 0.131 |
| 4 | 1.218 | 0.227 | 0.697 | 0.597 | 0.809 | 0.437 | 0.203 | 0.777 | 0.338 | 0.423 |
| 5 | 1.208 | 0.952 | 0.32 | 1.041 | 1.11 | 0.133 | 0.209 | 0.29 | 0.661 | 0.209 |
| 6 | 1.304 | 1.132 | 0.667 | 0.645 | 0.725 | 0.217 | 0.106 | 0.255 | 0.198 | 0.901 |

Table 4.3.23 cont.

## 1

XSA population numbers (Thousands)

|  | AGE |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 |

2000 2.85E+04 2.44E+04 1.86E+04 7.72E+03 3.23E+03 2.27E+03 $20018.09 \mathrm{E}+04 \quad 1.59 \mathrm{E}+04 \quad 8.58 \mathrm{E}+03 \quad 5.55 \mathrm{E}+03 \quad 1.87 \mathrm{E}+03 \quad 7.90 \mathrm{E}+02$
2002 1.10E $+05 \quad 5.93 \mathrm{E}+04 \quad 1.12 \mathrm{E}+04 \quad 5.38 \mathrm{E}+03 \quad 3.62 \mathrm{E}+03 \quad 5.91 \mathrm{E}+02$
2003 5.15E+04 7.78E+04 3.85E+04 5.07E+03 2.20E+03 2.15E+03
$2004 \quad 1.43 \mathrm{E}+04 \quad 3.62 \mathrm{E}+04 \quad 4.23 \mathrm{E}+04 \quad 1.84 \mathrm{E}+04 \quad 2.28 \mathrm{E}+03 \quad 6.35 \mathrm{E}+02$
$2005 \quad 1.82 \mathrm{E}+04 \quad 1.09 \mathrm{E}+04 \quad 2.59 \mathrm{E}+04 \quad 2.54 \mathrm{E}+04 \quad 6.70 \mathrm{E}+03 \quad 6.16 \mathrm{E}+02$
2006 8.23E+04 1.39E+04 5.55E+03 1.45E+04 1.34E+04 $4.80 \mathrm{E}+03$
2007 2.12E+04 $6.48 \mathrm{E}+04 \quad 1.01 \mathrm{E}+04 \quad 3.07 \mathrm{E}+03 \quad 9.71 \mathrm{E}+03 \quad 8.90 \mathrm{E}+03$
2008 1.09E+04 1.60E+04 $4.41 \mathrm{E}+04 \quad 4.98 \mathrm{E}+03 \quad 1.16 \mathrm{E}+03 \quad 5.94 \mathrm{E}+03$
$20096.65 \mathrm{E}+03 \quad 8.46 \mathrm{E}+03 \quad 1.11 \mathrm{E}+04 \quad 2.58 \mathrm{E}+04 \quad 2.91 \mathrm{E}+03 \quad 4.89 \mathrm{E}+02$
Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+00 \quad 5.27 \mathrm{E}+03 \quad 6.32 \mathrm{E}+03 \quad 8.01 \mathrm{E}+03 \quad 1.39 \mathrm{E}+04 \quad 1.98 \mathrm{E}+03$
Taper weighted geometric mean of the VPA populations:
$4.63 \mathrm{E}+04 \quad 3.75 \mathrm{E}+04 \quad 2.30 \mathrm{E}+04 \quad 1.11 \mathrm{E}+04 \quad 4.64 \mathrm{E}+03 \quad 1.93 \mathrm{E}+03$
Standard error of the weighted Log(VPA populations) :

|  | 0.8688 | 0.7471 | 0.6533 | 0.6353 | 0.69 | 0.9267 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Log catchability residuals.

Fleet: SCOGFS

| Age |  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | -0.33 | 0.26 | 0.01 | -0.03 | 0.2 | 0.34 | -0.19 | 99.99 | 0.3 |  |
|  | 2 | -0.44 | 0.54 | 0.46 | 0.01 | 0.22 | 0.32 | -0.29 | 99.99 | -0.24 |  |
|  | 3 | -0.38 | 0.27 | 0.37 | 0.23 | 0.2 | 0.01 | -0.55 | 99.99 | -0.14 |  |
|  | 4 | -0.17 | 0.62 | 0.4 | 0.49 | 0.83 | 0 | -1.11 | 99.99 | -0.28 |  |
|  | 5 | -0.13 | 0.29 | 0.73 | -0.36 | 1.03 | 0.18 | -0.54 | 99.99 | -0.13 |  |
|  | 6 | 0.07 | 0.21 | 0 | -0.09 | 0.14 | -0.14 | -0.34 | 99.99 | -0.07 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 | 99.99 | -0.55 | -0.21 | 0.11 | 99.99 | 0.57 | -0.04 | 0.16 | -0.13 | -0.48 |
|  | 2 | 99.99 | -0.57 | -0.71 | 0.2 | 99.99 | 0.39 | 0.71 | -0.17 | -0.12 | -0.32 |
|  | 3 | 99.99 | -0.04 | -0.48 | -0.27 | 99.99 | -0.01 | 0.39 | 0.36 | 0.05 | -0.01 |
|  | 4 | 99.99 | -0.78 | -0.81 | -0.54 | 99.99 | 0.46 | 0.45 | 0.69 | -0.43 | 0.19 |
|  | 5 | 99.99 | -0.29 | -0.89 | 0.52 | 99.99 | -0.37 | 0.98 | 0.18 | 0.03 | -1.23 |
|  | 6 | 99.99 | -0.33 | -0.01 | 0.29 | 99.99 | 0.13 | 0.36 | -0.1 | 0 | -0.46 |

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 4 | 5 | 6 |
| :---: | ---: | ---: | ---: |
| Mean Log | -2.4997 | -2.7645 | -2.7645 |
| S.E(Log q | 0.605 | 0.6314 | 0.2276 |

Regression statistics :
Ages with q dependent on year class strength

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Mean Log q

Ages with $q$ independent of year class strength and constant w.r.t. time.

Table 4.3.23 cont.

Terminal year survivor and F summaries :
Age 1 Catchability dependent on age and year class strength
Year class $=2008$

| Fleet | Int |  | Ext | VarRatio | N | Scaled |  | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | s.e |  |  |  |  | Weights |  |
| SCOGFS | 3233 | 0.393 | 0 | 0 |  | 1 | 0.69 | 0.06 |
| P shrink: | 37472 | 0.75 |  |  |  |  | 0.199 | 0.005 |
| F shrinke | 3298 | 1 |  |  |  |  | 0.111 | 0.059 |


| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 5273 | 0.33 | 0.77 |  | 3 | 2.349 | 0.038 |

1
Age 2 Catchability dependent on age and year class strength
Year class $=2007$

| Fleet | Int |  | Ext | Var | N | Scaled Weights |  | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | s.e | Ratio |  |  |  |  |
| SCOGFS | 5178 | 0.291 | 0.092 | 0.32 |  | 2 | 0.756 | 0.117 |
| P shrink: | 23026 | 0.65 |  |  |  |  | 0.171 | 0.027 |
| F shrinke | 2381 | 1 |  |  |  |  | 0.073 | 0.238 |


| Survivors | Int | Ext | N |  | Var | F |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of ) | s.e | s.e |  | Ratio |  |  |
| 6316 | 0.26 | 0.38 |  | 4 | 1.472 | 0.097 |

Age 3 Catchability dependent on age and year class strength
Year class $=2006$


| Survivors | Int | Ext | N |  | Var | F |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of J | s.e | s.e |  |  | Ratio |  |
| 8010 |  | 0.2 | 0.16 |  | 5 | 0.804 |

Table 4.3.23 cont.

Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2005$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SCOGFS | 13927 | 0.2 | 0.061 | 0.3 | 4 | 0.919 | 0.421 |
| F shrinke | 13102 | 1 |  |  |  | 0.081 | 0.442 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of ) | s.e | s.e |  | Ratio |  |  |  |
| 13859 | 0.2 | 0.05 | 5 | 0.256 | 0.423 |  |  |

Age 5 Catchability constant w.r.t. time and dependent on age

## Year class $=2004$



Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 1980 | 0.2 | 0.3 |  | 6 | 1.505 | 0.209 |

Age 6 Catchability constant w.r.t. time and age (fixed at the value for age) 5
Year class $=2003$


Weighted prediction :


Table 4.3.24. Haddock in VIb. Final runs with old survey indices. Fishing mortality+at+age.

Run title : HADDOCK LANDISC 2010 ROCKALL At 4/06/2010 9:34

Terminal Fs derived using XSA (With F shrinkage)

| Table 8 | Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.238 | 0.1758 | 0.1048 | 0.14 | 0.0507 | 0.2401 | 0.1661 | 0.2422 | 0.4954 |  |
| 2 | 0.5889 | 0.4825 | 0.3323 | 0.2864 | 0.2495 | 0.5699 | 0.3407 | 0.5778 | 0.7266 |  |
| 3 | 0.8877 | 0.7718 | 0.6442 | 0.5871 | 0.5516 | 0.4849 | 0.3129 | 0.6515 | 0.8882 |  |
| 4 | 0.9072 | 0.9576 | 0.4742 | 0.8803 | 0.7608 | 0.5114 | 0.3373 | 0.5054 | 0.9322 |  |
| 5 | 0.3655 | 0.9037 | 0.8954 | 0.458 | 0.7287 | 0.6364 | 0.5523 | 0.5926 | 0.8594 |  |
| 6 | 0.5353 | 0.2404 | 0.6347 | 0.7356 | 0.2589 | 0.5093 | 0.8655 | 0.6894 | 1.2334 |  |
| +gp | 0.5353 | 0.2404 | 0.6347 | 0.7356 | 0.2589 | 0.5093 | 0.8655 | 0.6894 | 1.2334 |  |
| 0 FBAR 2 | 0.6873 | 0.7789 | 0.5865 | 0.553 | 0.5726 | 0.5506 | 0.3858 | 0.5818 | 0.8516 |  |


| Table 8 Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.3847 | 0.1106 | 0.1445 | 0.1521 | 0.0746 | 0.0667 | 0.0394 | 0.0809 | 0.0496 | 0.0377 | 0.0561 |
|  | 2 | 0.8444 | 0.1469 | 0.2321 | 0.4085 | 0.134 | 0.4742 | 0.1245 | 0.1835 | 0.1653 | 0.097 | 0.1486 |
|  | 3 | 1.0099 | 0.2663 | 0.593 | 0.5409 | 0.3126 | 0.38 | 0.3911 | 0.504 | 0.3373 | 0.1313 | 0.3242 |
|  | 4 | 1.218 | 0.2267 | 0.6971 | 0.5972 | 0.8087 | 0.4371 | 0.203 | 0.7769 | 0.3378 | 0.4226 | 0.5125 |
|  | 5 | 1.2083 | 0.9516 | 0.3204 | 1.0406 | 1.1101 | 0.1326 | 0.2093 | 0.2903 | 0.6613 | 0.2086 | 0.3867 |
|  | 6 | 1.3039 | 1.1315 | 0.6673 | 0.6451 | 0.7252 | 0.2174 | 0.1057 | 0.2554 | 0.1975 | 0.901 | 0.4513 |
|  | +gp | 1.3039 | 1.1315 | 0.6673 | 0.6451 | 0.7252 | 0.2174 | 0.1057 | 0.2554 | 0.1975 | 0.901 |  |
| 0 | FBAR 2 | 1.0702 | 0.3979 | 0.4606 | 0.6468 | 0.5914 | 0.356 | 0.232 | 0.4387 | 0.3754 | 0.2149 |  |

Table 4.3.25 Haddock in VIb. Final runs with old survey indices. Stock number ( ${ }^{*} 10^{3}$ )-at-age.

Run title : HADDOCK LANDISC 2010 ROCKALL
At 4/06/2010 9:34
Terminal Fs derived using XSA (With F shrinkage)


Table 4.3.26 Haddock in VIb. Final run with old survey indices. Summary table.

## Run title : HADDOCK LANDISC 2004 ROCKALL

```
    At 4/06/2010 9:34
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    Table 16 Summary (without SOP correction)
    Terminal Fs derived using XSA (With F shrinkage)

| RE |  |  |  |  |  | TOTALE |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| Age 1 | TOTSPI | LANDIN | YIELD/S | FBAR 2- |  |  |
| 1991 | 110559 | 52117 | 16245 | 5655 | 0.3481 | 0.6873 |
| 1992 | 110271 | 51603 | 19884 | 5320 | 0.2675 | 0.7789 |
| 1993 | 124127 | 55617 | 20589 | 4784 | 0.2324 | 0.5865 |
| 1994 | 69118 | 56880 | 24983 | 5733 | 0.2295 | 0.553 |
| 1995 | 61492 | 48391 | 30213 | 5587 | 0.1849 | 0.5726 |
| 1996 | 62624 | 47827 | 26005 | 7075 | 0.2721 | 0.5506 |
| 1997 | 71921 | 41803 | 22343 | 5166 | 0.2312 | 0.3858 |
| 1998 | 73089 | 44105 | 21331 | 4984 | 0.2336 | 0.5818 |
| 1999 | 48886 | 33219 | 16703 | 5221 | 0.3126 | 0.8516 |
| 2000 | 28453 | 23341 | 11867 | 4558 | 0.3841 | 1.0702 |
| 2001 | 80937 | 21797 | 6957 | 1918 | 0.2757 | 0.3979 |
| 2002 | 109783 | 36340 | 7340 | 2571 | 0.3503 | 0.4606 |
| 2003 | 51514 | 37897 | 14224 | 5961 | 0.4191 | 0.6468 |
| 2004 | 14339 | 27162 | 17881 | 6400 | 0.3579 | 0.5914 |
| 2005 | 18178 | 22484 | 18266 | 5191 | 0.2842 | 0.356 |
| 2006 | 82314 | 26625 | 16213 | 2759 | 0.1702 | 0.232 |
| 2007 | 21196 | 28290 | 13823 | 3348 | 0.2422 | 0.4387 |
| 2008 | 10855 | 30998 | 24981 | 4205 | 0.1683 | 0.3754 |
| 2009 | 6648 | 21056 | 16972 | 3173 | 0.187 | 0.2149 |
|  |  |  |  |  |  |  |
| Arith. |  |  |  |  |  |  |
| Mean | 60858 | 37240 | 18254 | 4716 | 0.2711 | 0.5438 |
| OUnits | (Thousai | (Tonnes | (Tonnes | (Tonnes) |  |  |

Table 4.3.27 Haddock in VIb. Input data to short-term forecast (Data from final run with old survey indices).

MFDP version 1a
Run: had10
Time and date: 19:08 04,06,2010
Fbar age range (Total) : 2-5
Fbar age range Fleet $1: 2-5$

| 2010 |  |  | M | Mat |  | P | PM | SWt |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |  |  |  |
|  | 1 | 7181 |  | 0.2 |  | 0 | 0 | 0 | 0.187 |
|  | 2 | 5273 |  | 0.2 |  | 0 | 0 | 0 | 0.239 |
|  | 3 | 6316 |  | 0.2 |  | 1 | 0 | 0 | 0.299 |
|  | 4 | 8010 |  | 0.2 |  | 1 | 0 | 0 | 0.374 |
|  | 5 | 13859 |  | 0.2 |  | 1 | 0 | 0 | 0.542 |
|  | 6 | 1980 |  | 0.2 |  | 1 | 0 | 0 | 0.56 |
|  | 7 | 470 |  | 0.2 |  | 1 | 0 | 0 | 0.836 |




| Catch Age | Sel |  | CWt | DSel | DCWt |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.0112 | 0.173 | 0.0449 | 0.188 |
|  | 2 | 0.0496 | 0.26 | 0.099 | 0.243 |
|  | 3 | 0.2113 | 0.332 | 0.1129 | 0.279 |
|  | 4 | 0.4041 | 0.394 | 0.1084 | 0.315 |
|  | 5 | 0.326 | 0.575 | 0.0607 | 0.318 |
|  | 6 | 0.3963 | 0.595 | 0.055 | 0.433 |
|  | 7 | 0.4274 | 0.858 | 0.0239 | 0.464 |

Input units are thousands and kg - output in tonnes

Table 4.3.28. Haddock in VIb. Short-term forecast.
MFDP version 1a
Run: had10
Time and date: 19:08 04,06,2010
Fbar age range (Total) : 2-5
Fbar age range Fleet $1: 2-5$


Input units are thousands and kg - output in tonnes

Table 4.3.29. Haddock in VIb. Detailed short-term forecast output.


Input units are thousands and kg - output in tonnes

Table 4.3.30. Haddock VIb. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

| Year-class |  |  | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock No. (thousands) |  |  | 21196 | 10855 | 6648 | 7181 | 56520 |
| of |  | year-olds |  |  |  |  |  |
| Source |  |  | XSA | XSA | XSA | RCT3 | GM91-07 |
| Status Quo F: |  |  |  |  |  |  |  |
| \% in | 2010 | landings | 24.2 | 9.2 | 1.5 | 0.3 |  |
| \% in | 2011 | landings | 18.4 | 14.1 | 6.7 | 2.0 | 3.2 |
| \% in | 2010 | SSB | 21.6 | 13.6 | 0.0 | 0.0 | - |
| \% in | 2011 | SSB | 21.2 | 13.9 | 11.1 | 0.0 | 0.0 |
| \% in | 2012 | SSB | 15.0 | 12.2 | 10.1 | 14.4 | 0.0 |

GM : geometric mean recruitment

$$
\text { Haddock } V \mid b: Y e a r-c l a s s \% \text { contribution to }
$$



Figure 4.3.1. Length distribution and quantity of haddock lifted onboard and landings by Scottish trawlers in 1999 and 2001 (unpublished data, Newton, 2004).


Figure 4.3.2. Distribution of haddock (catch per 30 minutes) on the Rockall Bank in 1995-1999 and 2008-2009 from the Scottish trawl survey.


Figure 4.3.3. Rockall haddock in VIb. Scottish, Irish and Russian effort in 1985-2009.


Figure 4.3.4. Lpue and cpue of the fleets fishing for Rockall haddock. Note that Scottish and Irish effort data are not reliable because reporting is not mandatory.

1 - Scottish lpue (all gears)
2 - Irish trawlers lpue
3 - Cpue of Russian trawlers (BMRT type, tonnage class 10 in 1999-2007, and tonnage class 9 in 20082009).


Figure 4.3.5. Dynamics of haddock total biomass (ICES, 2008a; ICES, 2008b) and directed fishing efficiency ( $\mathbf{t}$ per a trawling hour) for tonnage class 10 vessels in 1999-2007.


Figure 4.3.6. Total landings and discards of Rockall haddock ('000 individuals).


Figure 4.3.7. Total landings and discards of Rockall haddock (tonnes).



Figure 4.3.8. Haddock in VIb. Mean weights-at-age a) in catch and b) in stock.


Figure 4.3.9. Haddock in VIb. Log catch (with discards in numbers)-at-age by year.


Figure 4.3.10. Haddock in VIb. Log landings (in numbers)-at-age by year.


Figure 4.3.11. Haddock in VIb. Log catch (with discards, in numbers)-at-age by year class.


Figure 4.3.12. Haddock in VIb. Log landings (without registered discards, in numbers)-at-age by year class.


Figure 4.3.13. Haddock in VIb. Catch curves (with registered discards).


Figure 4.3.14. Haddock in VIb. Catch curves (landings without registered discards).


Figure 4.3.15. Haddock in VIb. Log survey cpue at age by year.


Figure 4.3.16. Haddock in VIb. Log survey cpue by year class.


Figure 4.3.17. Haddock in VIb. Log survey cpue at age.


Figure 4.3.18. SURBA analysis for Rockall haddock.


Mean total mortality


Total stock biomass


Spawning stock biomass


Recruitment


Figure 4.3.18a. SURBA analysis for Rockall haddock. Retrospective plots.

SCOGFS: Comparative scatterplots at age


Figure 4.3.18b. SURBA analysis for Rockall haddock. Pairwise plots of age.



Figure 4.3.19. Haddock in VIb. Log catchability residual plots (shrinkage 1.0). XSA run: catchability dependent on stock size at ages $<4$. XSA run to 2008 . Old survey indices data.


Figure 4.3.20. Haddock in VIb. Log catchability residual plots (shrinkage 1.0). XSA run: catchability dependent on stock size at ages $<4$. XSA run to 2008. Standardized survey indices data.


Figure 4.3.21. Haddock in VIb. Log catchability residual plots (shrinkage 1.0). Final XSA: catchability dependent on stock size at ages <4. XSA run to 2009. Standardized survey indices data.


Figure 4.3.22. Haddock in VIb. Log catchability residual plots (shrinkage 1.0). Final XSA: catchability dependent on stock size at ages $<4$. XSA run to 2009 . Old survey indices data.




Figure 4.3.23. Haddock in VIb. Comparison of the final assessment (in red) with the assessment in which were used standardized survey indices (in blue).


Figure 4.3.24. Haddock in VIb. Adjusted Scottish groundfish survey cpue from the final XSA run plotted against VPA numbers (shrinkage 1.0)-at-age. Catchability dependent on stock size at ages <4.







Figure 4.3.25. Haddock in VIb. Survey indices and XSA estimates (shrinkage 1.0)-at-age. Final XSA: catchability dependent on stock size at ages $<4$.



Fis hing mortality (F 2-5)


Figure 4.3.26. Haddock in VIb. Retrospective analyses (F shrinkage 1.0).


Figure 4.3.27. Haddock in VIb. $F$ at age ( F shrinkage 1.0).


Figure 4.3.28. Haddock in VIb. XSA and SURBA analyses.





Figure 4.3.29. Haddock in VIb. Summary plots.


Figure 4.3.30. Haddock in VIb. Comparison of the current assessment (in red) with the previous one (in black).


Figure 4.3.31. Haddock in VIb. Comparison observed and predicted by StatCam survey index and catch biomass . Scenario 2.


Figure 4.3.32. Haddock in VIb. Log catchability residuals plot for survey biomass index. Scenario 2 of Statcam run.


Figure 4.3.33. Haddock in VIb. Population biomass, SSB, fishin mortality and recruitment by Statcam estimation. Scenario 2.


Figure 4.3.34. Haddock in VIb. Comparison of VPA assessment with the statistical catch-at-age model StatCam assessment.


Figure 4.3.35. Haddock in VIb. Scottish Groundfish survey indices of haddock at age 0.


Figure 4.3.36. Haddock in VIb. VPA numbers-at-age 1 from XSA plotted against Scottish Groundfish survey indices of haddock at age 0 .


Figure 4.3.37. Haddock in Division VI b. Discard proportion-at-age by year and mean discard proportion-at-age for two periods, 1991-2009 and 1999-2009.

Figure Haddock, Rockall. Short term forecast



Figure 4.3.38. Haddock in VIb. Short-term forecast.

Figure Haddock, Rockall. Sensitivity analysis of dhort tem forecact.


Figure 4.3.39. Haddock in VIb. Delta plots from selectivity analysis.

Figure Haddock, Rockall. Probability profiles for short term forecast.


## 

Figure 4.3.40. Haddock in VIb. Probability plots for yield in 2011 and SSB in 2012.


Figure 4.3.41. Haddock VIb. Medium-term analysis.


Figure 4.3.42. Haddock VIb. Medium-term analysis.

Rockall Haddock


Figure 4.3.43. Haddock in VIb. Biological reference points.

## Rockall Haddock: Stock and Recruitment



Figure 4.3.44. Haddock in VIb. SSB and recruitment.

## Rockall Haddock: Field per Recruit



Figure 4.3.45. Haddock in VIb. Yield-per-recruit.


Figure 4.3.46. Haddock in VIb. Fitted stock-recruit relationships with $\mathbf{1 0 0 0}$ MCMC re-samples. The left-hand plots show the deterministic fit (blue) as well as the confidence intervals from converged estimates of $\mathrm{F}_{\mathrm{msy}}$ (red). Right-hand panels show the fits from the first 100 converged MCMC re-samples for illustration. The legends show the number of converged values for $\mathrm{F}_{\text {msy }}$ from $\mathbf{1 0 0 0}$ re-samples.


Figure 4.3.47. Haddock in VIb. Estimates of $F$ reference points and equilibrium yield and SSB against mortality using a Beverton and Holt recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical.

## had10 Smooth hockeystick



Figure 4.3.48. Haddock in VIb. Estimates of F reference points and equilibrium yield and SSB against mortality using a hockey stick recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical.


Figure 4.3.49. Estimates of F reference points and equilibrium yield and SSB against mortality using a Ricker recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical.
had10 - Per recruit statistics


Figure 4.3.50. Fitted of F reference points and equilibrium yield and SSB. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals (red) and the right hand plots show the fit for the first 100 iterations. The top two plots are identical.

### 5.1 Northern Shelf overview

There is no overview.

### 5.2 Anglerfish (Lophius piscatorius and L. budegassa) in Division IIa, IIIa, Subarea IV and VI

The WGNSDS considered the stock structure of anglerfish on a wider European scale in 2004, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division IIa. However, for the purposes of reporting, anglerfish in IIa is treated in a separate section (5.2.2) from anglerfish on the northern shelf (Div. IIIa, Subarea IV and VI, Section 5.2.1), but the advice refers to both.

### 5.2.1 Anglerfish in Division IIIa, Subarea IV and VI

There has been no assessment of the anglerfish stock on the northern shelf since 2003. Recent ACFM review groups have highlighted the generally poor data for this stock and the need to continue with the recently instigated data collection schemes (both survey and commercial data) in order to obtain time-series of sufficient length. Since 2005, an annual science- industry partnership survey has been conducted by the Scottish, and in some years, Irish institutes: updates to these survey data are presented this year, along with updates to catch and effort data where available.

## ICES advice applicable to 2009 and 2010

The ICES advice for 2009 (Single Stock Exploitation Boundaries) was as follows:
"The new data available for this stock do not change the perception of the stock and do not give reason to change the advice from 2007. The advice for the fishery in 2009 is therefore the same as the advice given in 2007 for the 2008 fishery: The effort in fisheries that catch anglerfish should not be allowed to increase and the fishery must be accompanied by mandatory programmes to collect catch and effort data on both target and bycatch fish.
In addition, ICES offers the following considerations: Following ICES suggestions in 2005 a number of initiatives were instigated covering anglerfish in Division IVa and Subarea VI:

- dedicated Scottish and Irish scientific anglerfish surveys which are coordinated to involve the use of both research vessels and commercial fishing vessels;
- a Scottish tallybook scheme (linked to a longer time-series of personal diaries);
- increased observer coverage (short-term initiative in 2006).

Data are currently being gathered, with improvements to both industry-related data and surveys covering Subarea VI and part of the North Sea. There are currently 3 years of surveyderived absolute abundance estimates and 2 complete years of Scottish tallybook data providing commercial catch data."

The ICES advice for 2010 (Single Stock Exploitation Boundaries) was as follows:
"ICES advises on the basis of precautionary considerations that the effort in fisheries that catch anglerfish should not be allowed to increase."

### 5.2.1.1 General

## Stock description and management units

For the purposes of this section, the anglerfish stock on the Northern Shelf is considered to occur in Divisions IIa, IIIa (Skagerrak and Kattegat), Subarea IV (the North Sea) and Subarea VI (West of Scotland plus Rockall). Anglerfish in the North Sea and

Skagerrak/Kattegat were considered by this Working Group for the first time in 1999. In 2004, the WG was asked to consider the stock structure of anglerfish on a wider Northern European scale and despite a lack of conclusive evidence to indicate a single stock, anglerfish in IIa was included in the ToR at subsequent WG meetings.

Management of Northern Shelf anglerfish is based on separate TACs for the North Sea area and West of Scotland area. The following Table summarises ICES advice and actual management applicable for Northern Shelf anglerfish during 2003-2008.

|  | Single <br> stock <br> exploitation boundary | Basis | West of Scotland |  |  | North Sea |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  | TAC ${ }^{4}$ | \% change <br> in F <br> associated <br> with TAC | WG <br> landings | TAC ${ }^{5}$ | \% change <br> in F <br> associated <br> with TAC | WG <br> landings |
| 2003 | <67001) | Reduce F below $\mathrm{Fpa}_{\mathrm{pa}}$ | 3180 | $49 \%$ <br> reduction | 4126 | 7000 | $49 \%$ <br> reduction | 8268 |
| 2004 | <88002) | Reduce F below $\mathrm{F}_{\mathrm{pa}}$ 2) | 3180 | $48 \%$ <br> reduction | 3296 | 7000 | $48 \%$ <br> reduction | 9027 |
| 2005 | - | No effort increase ${ }^{2)}$ | 4686 | - | $\mathrm{n} / \mathrm{a}$ | 10314 | - | $\mathrm{n} / \mathrm{a}$ |
| 2006 | - | No effort increase ${ }^{2)}$ | 4686 | - | n/a | 10314 | - | $\mathrm{n} / \mathrm{a}$ |
| 2007 | - | No effort increase ${ }^{2)}$ | 5155 | - | $\mathrm{n} / \mathrm{a}$ | 11345 | - | $\mathrm{n} / \mathrm{a}$ |
| 2008 | - | No effort increase ${ }^{3)}$ | 5155 | - |  | 11345 | - |  |
| 2009 | - | No effort increase ${ }^{3)}$ | 5567 | - |  | 11345 | - |  |
| 2010 | - | No effort increase ${ }^{3)}$ | 5567 | - |  | 11345 | - |  |

All values in tonnes.
${ }^{1)}$ Advice for Division IIIa, Subarea IV and Subarea VIa combined.
${ }^{2)}$ Advice for Division IIIa, Subarea IV and Subarea VI combined.
${ }^{3)}$ Advice for Division IIa, Division IIIa, Subarea IV and Subarea VI combined.
${ }^{4)}$ TAC applies to $\mathrm{Vb}(E C)$, VI, XII and XIV.
${ }^{\text {5) }}$ TAC applies to IIa \& IV (EC)
Although there is no minimum landing size for this species, there is an EU minimum weight of 500 g for marketing purposes (EC Regulation 2406/96).

An additional quota of 1540 t is also available for EU vessels fishing in the Norwegian zone of Subarea IV in 2010.

## The fishery in 2009

A description of the fisheries on the northern shelf is given in Section 5.1 above.
UK (Scottish) vessels account for $47 \%$ of the reported anglerfish landings from the Northern Shelf area. The Danish and Norwegian fleets are the next most important exploiters of this stock in the North Sea while French vessels take the majority of the landings from the West of Scotland followed by the UK and Ireland.

The official landings by area are given in Table 5.2 .1 and the breakdown by country in Tables 5.2.2-4. In 2009, total [officially reported] landings ( 16539 t) were lower than in 2008 (17 300). This was largely due to a reduction in [officially reported] landings in Division VIa and IVa by the UK (Table 5.2.2). Total officially reported landings of anglerfish from the Northern Shelf are shown in Figure 5.2.1. During the 1970s landings were fairly stable at around 9000 t , but from about 1983 they increased steadily to a peak of over 35000 t in 1996, and then declined rapidly during the following five years. However, any subsequent declines in reported landings may have been due to restrictive TACs and are not necessarily representative of actual landings. The overall trend in landings is driven by the landings from the Northern North Sea and West of Scotland. Together these two areas account on average for approximately $80 \%$ of the total landings over 1973-2009.

Uptake of EC quota, based on the officially reported landings was as follows:

|  | TAC ${ }^{1}$ | Landings | Uptake (\%) | TAC |  |  | Landings | Uptake (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | VI | VI |  | IV <br> (Norwegian) | IIa \& IV | IIa \& IV <br> (total) | IIa \& IV <br> (total) |  |
| Belgium | 200 | 0 | 0 | 47 | 401 | 448 | 139 | 31 |
| Denmark |  | 0 |  | 1189 | 884 | 2073 | 1693 | 82 |
| France | 2462 | 2289 | 93 |  | 82 | 82 | 0 | 0 |
| Germany | 228 | 211 | 93 | 19 | 432 | 451 | 233 | 52 |
| Ireland | 557 | $419{ }^{2}$ | 75 |  |  |  | 0 |  |
| Netherlands | 193 | 0 | 0 | 17 | 303 | 320 | 53 | 17 |
| Spain | 214 | 0 | 0 |  |  |  | 0 |  |
| Sweden |  | 0 |  |  | 10 | 10 | 27 | 270 |
| UK (total) | 1713 | 2065 | 121 | 278 | 9233 | 9511 | 8172 | 86 |
| Total | 5567 | 4936 | 89 | 1550 | 11345 | 12895 | 10317 | 80 |

${ }^{1} \mathrm{TAC}$ applies to $\mathrm{VI}, \mathrm{Vb}(\mathrm{EC})$, and international waters of XII and XIV.
${ }^{2}$ Provisional
Catches in Division IIIa are not regulated: Table 5.2.4 shows the official landings which came to a total of 548 t in 2009. The landings by fleet for Denmark and Norway are given in Figures 5.2.2 and 5.2.3 respectively. The Irish fleet is dominated by demersal trawlers and so it is not shown here.

### 5.2.1.2 Data

## Landings

The TACs for both the West of Scotland and North Sea areas were reduced substantially in 2003 and 2004, and at previous WGs it has been highlighted that these reductions would likely imply an increased incentive to mis-report landings and increase discarding unless fishing effort was reduced accordingly (Section 6.4.6, ICES WGNSDS 2003). Anecdotal information from the fishery in 2003 to 2005 appeared to suggest that the TACs were particularly restrictive in these years. The official statistics for these years are, therefore, likely to be particularly unrepresentative of actual landings. The introduction of UK \& Irish legislation requiring registration of all fish buyers and sellers (See Section 1.7) may mean that the total reported landings from 2006 onwards are more representative of actual total landings.

In the meantime, collation of an international landings-at-age dataset is being hampered by the different approaches to age determination by the institutes which could provide these data. It has been proposed by ICES PGCCDBS that this be addressed using an anglerfish ageing exchange to be held in 2011.

The absence of a TAC for Subarea IV prior to 1999, means that before 1999, landings in excess of the TAC in other areas were likely to be misreported into the North Sea. In 1999, a precautionary TAC was introduced for North Sea anglerfish, but unfortunately for current and future reporting purposes, the TAC was set in accord with recent catch levels from the North Sea which includes a substantial amount misreported from Subarea VI. The area misreporting practices have thus become institutionalised and the statistical rectangles immediately east of the $4^{\circ} \mathrm{W}$ boundary (E6 squares) have accounted for a disproportionate part of the combined VIa/North Sea catches of anglerfish. The Working Group historically (prior to 2005) provided estimates of the actual Division VIa landings by adjusting the reported data for Division VIa to include a proportion of the landings declared from Division IVa in the E6 ICES statistical rectangles. This adjustment has been adapted to include landings declared from the whole of Area VI. Details of how the correction has been applied are given in the Stock Annex. Scottish officially reported landings adjusted for area misreporting are shown along with landings from Ireland, Denmark, France and Norway in Figure 5.2.4. Due to ongoing technical problems associated with changes to the Marine Scotland database and lack of landings data provided to the Working Group by some of the major nations exploiting the fishery, WG estimates of the actual Division VIa and IVa landings have not been calculated for recent years (2005-2008).

The corrected spatial distribution of anglerfish landings shows a typical pattern, with most landings being taken from the area around Shetland and also the area to the west of Scotland close to the shelf edge. Some landings, associated with the Nephrops fishery, are taken from the Fladen ground in the middle of the northern North Sea. A substantial amount of landings were taken from Rockall. The spatial distribution of Danish landings shows the typical pattern of higher landings around the Norwegian deeps. The Irish fishery in 2008 landed principally from the west coast of Ireland and in the south of Division VIa, with some landings from Rockall.

Consideration should be given in future to examining the distribution of landings combined with vessel monitoring system (VMS) data, perhaps using a kilowatt fishing hours metric to produce spatial distributions of lpue.

## Commercial catch-effort data

## Scotland

Reliable effort data (in terms of hours fished) are not available from the Scottish trawl fleets due to changes in the practices of effort recording and non-mandatory recording of hours fished in recent years. Further details can be found in Section B4 of the Stock Annex and the Report of the 2000 WGNSSK (ICES, 2001). Effort data in terms of days fished are available from official logbooks and these data are presented by gear in the report of WGNSDS 2007. However, given the uncertainties associated with the official landings from the recent past, no attempt has been made to use these data to calculate an lpue series and they have not been updated this year.

Attempts have recently been made to obtain more reliable data on catch and effort from the Scottish anglerfish fishery. In 2005, an analysis of data collated from the personal diaries of Scottish skippers operating across the Northern Shelf was presented to this WG (ICES, 2006 and Bailey et al., 2004). Following recommendations
made by ACFM that this data collection scheme should be continued and extended, in 2006, Marine Scotland Science (in consultation with the fishing industry) established a monkfish tallybook project. A fuller description and analysis of these data can be found in the WGNSDS 2008 Report and Dobby et al. (2008). However, at present there are problems in the scheme in terms of falling participation levels (four vessels in 2008; two vessels in 2009): this is unlikely to give a representative picture of the fishery and so updates of these data are not included.

## Ireland

Trends in official landings, effort in hours fished) from the Irish otter trawl fleets (OTB) operating in Division VIa and VIb are shown in Table 5.2.7 and Figure 5.2.5. This fleet is responsible for the majority of the landings from the south of Division VIa. Landings and effort data from the other fleets (1995-2006) are available in the Stock Annex. The Irish lpues from logbooks are shown in Figure 5.2.5. The timeseries show increasing trends in (particularly) Division VIa in recent years. However, it is not clear whether such trends are indicative of stock trends as such increases in lpue could also be due to changes in targeting behaviour due to reductions in fishing opportunities for other species and changes in reporting practices.

## Denmark

Danish logbook data for anglerfish landings and corresponding effort by main fishery in the North Sea and IIIA for the period 2000-2009 are shown in Table 5.2.5. Figure 5.2.6 and Table 5.2.8 show the fluctuations in lpue for anglerfish in mixed demersal fisheries and the shrimp fishery (small meshed). Of particular relevance is the series for the mixed demersal trawl fisheries in the North Sea including Nephrops trawls as these are where most anglerfish is taken (Table 5.2.5). Note the upwards trend, especially from 2003 to 2004 for all fisheries and the subsequent stabilisation in lpue. A time-series, 1997-2008, corrected for increase in fishing power for the shrimp trawl lpue indicates a declining trend over the time-series (Figure 5.2.6). There has been an increase in overall effort in 2008 and 2009 (Table 5.2.5 B).

Anecdotal information from Danish fishermen suggests that this apparent levelling off in lpue is due to the TAC constraints on the Danish fishery in the Norwegian EEZ since 2005, which was not in evidence in previous years. Although catch rates are not declining, the TAC constraints and possible technological creep currently render it problematic to use these logbook based lpues as indicators of stock abundance.

## Norway

Available logbook data from Norwegian trawlers have been examined for the possibility of establishing a cpue time-series for anglerfish. However, several problems were encountered in the dataset, and it is still considered insufficient for providing any reliable information on trends in stock abundance.

Six gillnetters have been included in a self-sampling scheme established along the Norwegian coast within IVa and IIIa. Detailed information about effort and catch will be provided through this scheme, and will potentially be valuable in future assessments of anglerfish in this area.

## Other countries

No effort data were available for the Spanish and French fleets operating in Subarea VI.

## Research vessel surveys

At previous meetings of this WG it has been concluded that the traditional groundfish surveys are ineffective at catching anglerfish and do not provide a reliable indication of stock size. As a result of this conclusion, and the urgent requirement for fishery-independent data, Marine Scotland Science, began a new joint science/industry survey in 2005. This is a targeted anglerfish survey using commercial gear. In 2006, 2007 and 2009, Ireland also participated extending the anglerfish survey to cover the remaining part of VIa (from 54030' to 56039') and, in 2006 and 2007, into ICES Areas VIIb,c,j. Further details of the survey including information on design, sampling, gear and vessel were recently considered by ICES WKAGME and are available in ICES (2009).

Results from previous surveys, as described in previous Working Group reports, did not take into account certain errors in the estimation process. In addition to reporting the results of the 2009 surveys, new abundance and biomass estimates are now provided for the 2005-2009 surveys (summarised in Table 5.2.9) with the appropriate error and its propagation (see WD 5, Fernandes, 2010 and WD 6, Yuan et al., 2009). The estimates presented this year, represent the best available knowledge to date from the five surveys carried out (2005-2009) and as such they take into account the following factors:

1 ) herding of anglerfish by the trawl doors and sweeps;
2 ) escapes of fish under the trawl footrope;
3 ) anglerfish abundance and biomass in the southern part of Area VI not covered in 2005 and 2008;

4 ) visual counts of anglerfish in areas closed to trawling at Rockall;
5 ) variability due to
5.1) sampling
5.2 ) missing ages
5.3 ) herding (based on experimental data)
5.4 ) footrope escapes (based on experimental data)

The estimates currently do not take account of the following:
1 ) areas in the central and southern North Sea (eastern part of ICES Division IVa and all of IVb and IVc );

2 ) areas inaccessible to the trawl in Division VIa.
Methods to account for these factors are under development.
The 2009 survey took place in April: the sample locations for $(\mathrm{n}=206)$ are illustrated in Figure 5.2.7 as the number density (number per square kilometre) and Figure 5.2.8 as the weight density (kilograms per square kilometre) of anglerfish. The highest densities of anglerfish occurred close to the 200 m contour in the northern and western areas, including the northern North Sea (particularly by weight). Very high densities were found on the east coast of the Rockall plateau. The results of the survey are presented in Table 5.2.9. The total estimate for the whole northern shelf in 2009 was 35800 t with $95 \%$ confidence limits of 28600 to 44700 tonnes. The Relative Standard Errors for the Scottish components were $12.1 \%$ and $14.5 \%$ for abundance and biomass respectively for the Northern shelf. The incomplete survey in ICES Area IV gave a slightly lower biomass of $17100 t$ than the largely complete survey estimate in ICES Area VI of 18700 t . The estimates-at-age (Figure 5.2.9) indicate that despite
corrections for catchability, which largely affect the smaller, younger fish, there is still an issue with catchability which is unaccounted for. It should also be noted that ageing of anglerfish is still uncertain. The last angler (Lophius spp.) otolith exchange took place in 2001 and the last black-bellied angler (L. budegassa) otolith exchange took place in 2004. Landa et al. (2008), however, noted that previously used ageing criteria are not accurate. There is ongoing research to establish if a new protocol should be established when using illicia to estimate age. Full exchanges of otoliths and illicia are therefore recommended for 2011, when new ageing criteria are expected (ICES 2010).

The time-series estimates indicate that biomass increased in most areas from 2005 to 2008, but decreased in 2009, whereas abundance was stable in 2005-2007 and has since declined (Table 5.2.9 and Figure 5.2.10). The estimates of abundance of anglerfish from the surveys from 2005-2009 are in line with previous attempts to quantify their abundance (ICES 2004): the last assessment estimated the total stock biomass to be just under 37000 t in 2002. There are still several factors which make the survey estimates likely to be underestimates or minimum estimates. Firstly, although experiments have been carried out to estimate escapes from under the footrope, and a model applied to account for this component of catchability, the estimates of younger anglerfish (ages $0-3$ ) still look to be underestimated (Figure 5.2.9). This could be due to either a net selectivity issue, or an availability [to the trawl] issue, as it is known that younger fish occur in shallower water (Hislop et al., 2001). Methods to compensate for these additional catchability and availability factors are being considered by developing a survey based assessment model. Secondly, the area considered was not complete. Although only a small part of ICES Area VI was missed, quite a large part of ICES Area IV was not surveyed (Figure 5.2.8). Although repeated requests have been made to countries with an interest in the anglerfish fishery to consider participating, no other countries have done so, with the exception of the Irish who participated in 2006, 2007 and 2009. The problem is, therefore, being tackled by an examination of data from the International Bottom Trawl survey. If a relationship can be found between the IBTS survey data and the data from the anglerfish survey where they overlap, then abundance estimates in the southern North Sea could be derived by interpolation where there is only IBTS data. These methods are currently under development (see ICES WKAGME 2009).

### 5.2.1.3 Historical stock development

There has been no assessment of this stock since the length based assessment presented in ICES (2004). This indicated a total stock size of approximately 36590 t in 2002.

The estimates of abundance of anglerfish from the surveys from 2005-2009 are in line with these previous attempts to quantify their abundance. There are still several factors which make the survey estimates likely to be underestimates or minimum estimates (see above).

### 5.2.1.4 Short-term projections

In the absence of an age based assessment, there are no short-term projections for this stock.

The European Commission's Consultation on Fishing Opportunities for 2010 (COM(2009) 224), sets out an approach to set the TACs in cases where scientific advice on an appropriate catch level is provided, but a quantified stock assessment calculation is not available, usually for reasons of uncertain data quality.

| Subarea | Period | avg survey biomass <br> (tonnes) | $\%$ <br> change | TAC <br> change | 2010 TAC* | 2011 TAC |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- |
| IV | $2005-2007$ | 23,210 |  |  |  |  |
|  | $2008-2009$ | 23,365 | +1 | NA |  |  |
|  |  |  |  |  |  |  |
| VI | $2005-2007$ | 20,265 |  |  |  |  |
|  | $2008-2009$ | 21,192 | +5 | NA |  |  |
|  |  |  |  |  |  |  |
| VIa | $2005-2007$ | 12,548 |  |  |  |  |
|  | $2008-2009$ | 10,306 | -18 | NA |  |  |
|  |  |  |  |  |  |  |
| VIb | $2005-2007$ | 7,717 |  |  |  |  |
|  | $2008-2009$ | 10,886 | +41 | +15 | 3711 |  |

*2010 TACs are set according to ICES Subareas: the proportion of biomass in Subarea VI attributed to Division VIb was used to apportion the Subarea VI TAC into an allocation for VIb.

Anglerfish will come under this "Category 6 to 9 " grouping of stocks in 2010. The Table above shows the outcome of applying the Commission's rules 4 and 5 of Annex III for Category 6 to 9 based on the survey data from 2008 compared to the outcome (actual TAC in 2009).

In terms of setting the TAC for 2011, this needs to be based on the 2010 survey which has recently been completed: The data from the 2010 survey should be considered along with other ICES' survey updates later on in the year.

### 5.2.1.5 MSY evaluations

In terms of the status of $F$ in relation to $\mathrm{F}_{\text {msy }}$ there are two major uncertainties. The first is the value of $\mathrm{F}_{\mathrm{msy}}$. Previous WG have considered that the fishing mortality corresponding to $35 \%$ of the unfished SSB/R could be an approximation of $\mathrm{F}_{\text {MSY: }}$ : this is what $\mathrm{F}_{\mathrm{pa}}$ was set to $\left(\mathrm{F}_{35 \% \mathrm{SPR}}=\mathrm{F}_{\mathrm{pa}}=0.30\right)$. Another suitable proxy might be $\mathrm{F}_{0.1}$, which like $\mathbf{F}_{35 \% \text { SPR, }}$, would be derived from a yield-per-recruit analysis. However, as yet no assessment is available to determine the fishing mortality [selection] pattern which is required for a $\mathrm{Y} / \mathrm{R}$ analysis. The second uncertainty is the current level of fishing mortality, where, in the absence of an assessment, this is also unknown. However, if the ageing of anglerfish in the surveys described above is assumed to be accurate and the survey is sampling the population in an unbiased way then a provisional estimate of total mortality ( $Z$ ) from abundance curves would be approximately 0.6 . Given an assumed natural mortality of 0.15 (as used in past assessments) this would imply an F at about 0.45 . The last time a yield-per-recruit was carried out (ICES 2004), F $\mathrm{F}_{0.1}$ was estimated at 0.12 and $\mathrm{F}_{35 \% \text { SPR }}$ was 0.12 . $\mathrm{F}_{\mathrm{pa}}$ for this stock was based on an earlier estimate of $\mathrm{F}_{35 \% \text { SPR }}$ at 0.3 . Even with the various uncertainties expressed, it seems likely that this stock is, therefore, being exploited at a fishing mortality in excess of Fmsy.

### 5.2.1.6 Biological reference points

|  | Type | Value | Technical basis |
| :--- | :--- | :--- | :--- |
|  | $B_{\text {lim }}$ | Not <br> defined | There is currently no biological basis for defining Blim |
| Precautionary <br> approach | $\mathrm{B}_{\mathrm{pa}}$ | Not <br> defined |  |
|  | $\mathrm{F}_{\text {lim }}$ | Not <br> defined | There is currently no biological basis for defining Flim |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.30 | $\mathrm{F}_{35 \% \text { SPR }}=0.30$. This fishing mortality corresponds to 35\% <br> of the unfished SSB/R. It is considered to be an <br> approximation of FMSY. |
| Targets | $\mathrm{F}_{\mathrm{y}}$ | Not <br> defined |  |

(unchanged since 1998).

### 5.2.1.7 Management plans

There is no management plan for this stock.

### 5.2.1.8 Uncertainties and bias in assessment and forecast

This WG has previously attempted assessments of the anglerfish stock(s) within its remit using a number of different approaches. As yet none have proved entirely satisfactory. The catch-at-length analysis used in previous years appears to have addressed a number of the suspected problems with the data due to the rapid development of the fishery, and has also provided a satisfactory fit to the catch-atlength distribution data. However, since 2003, the WG has been unable to present an analytic assessment due to the lack of reliable fishery and insufficient survey information, and in addition it is not known to what extent the dynamic pool assumptions of the traditional assessment model are valid for anglerfish.

## Commercial data

For a number of years the WG has expressed concerns over the quality of the commercial catch-at-length data because of:

- Accuracy of landings statistics due to species and area misreporting.
- Lack of information on total catch and catch composition of gillnetters operating on the continental slope to the north west of the British Isles (See the Stock annex for further details of this fishery).

However, the introduction of legislation on buyers and sellers registration in the UK and Ireland since 2006 may mean that the reported landings for 2006 onwards are more reliable.

The recent Scottish tallybook scheme has been implemented as part of a long-term approach to provide better information on the fishery. Although the time-series of data is currently short, the scheme did have the potential to deliver relatively extensive information on spatial and depth distribution of catch rates provided that participation remains high. In addition to total catch rate information, the fishermen were also asked to provide information on landings by size category, discards, catches of mature females and bycatches of other species. However, participation in this scheme has fallen significantly and is now hampered by data sensitivities associ-
ated with the compliance of fishery regulations. The tally book programme is likely to be terminated as a result.

## Survey data

In addition to obtaining estimates of abundance from swept area methods (and in future a times-series of data for use in survey based assessments), a visual count method is being developed at Marine Scotland Science to provide alternative estimates of anglerfish density. It is also anticipated that the new Scottish-Irish science/industry survey will provide further useful information on the biology and stock structure of anglerfish. So far, a total of 48 live anglerfish have been tagged with data storage tags on the Marine Scotland Science surveys which if and when recovered will provide information on the vertical migration, depth distribution and temperature regime of individuals (recently, a tag has been returned from a fish which was tagged in 2005; these data are currently being retrieved from the tag manufacturer). Tagging carried out on the Irish survey (800 ribbon tags) should also provide information on movement of anglerfish.

In 2006, 2007 and 2009 Ireland extended the survey area to include the more southerly regions of the Northern Shelf stock of anglerfish area not covered by the Scottish survey. However the participation of other nations in a collaborative survey to include coverage of waters in the east and south of the North Sea would be invaluable.

## Biological information

Knowledge of the biology of anglerfish is improving. Some of the basic biological parameters used in the assessments, such as mean weight-at-age in the stock, are now becoming available from the industry science surveys. Difficulties still remain in finding mature females. However, recent studies by Laurenson et al. $(2005 ; 2008)$ carried out whilst observing the fishery, have obtained similar growth parameters and maturity ogives to those previously used. A further discussion of the biology can be found in the Stock Annex.

In addition, ageing has not been validated and should still be regarded as uncertain. An ageing exchange is due to be carried out in 2011.

## Stock structure

Currently, anglerfish on the Northern Shelf are split into Subarea VI (including $\mathrm{Vb}(\mathrm{EC})$, XII and XIV) and the North Sea (\& IIa (EC)) for management purposes. However, genetic studies have found no evidence of separate stocks over these two regions (including Rockall) and particle-tracking studies have indicated interchange of larvae between the two areas (Hislop et al., 2001). So, at previous WGs, assessments have been made for the whole Northern Shelf area combined. In fact, both microsatellite DNA analysis (O'Sullivan et al., 2005) and particle tracking studies carried out as part of EC 98/096 (Anon, 2001) also suggested that anglerfish from further south (Subarea VII) could also be part of the same stock.

Following the recent expansion of the anglerfish fishery in ICES Divisions IIa and V, in 2004 the WG group was asked to consider the stock structure on the wider Northern European scale (Section 16 of the WGNSDS 2004 Report). It was concluded that there was currently insufficient information to conclusively define new stock areas for assessment and further co-ordinated work is still required. Given the request to also assess anglerfish in Division IIa and that there may be an extension to include ICES Division V in the near future, the likely spatial disaggregation of the stock (drift of larvae and possible migration of mature fish back into deeper water) means that
any assessment model would need to be spatially structured, possibly supported by assessments for each of the stock units separately. Given the problems with data quality associated with Northern Shelf anglerfish, the WG wishes to highlight fundamentals required for a wider area assessment:

- Accurate information on the spatial distribution of catch and effort;
- Data on movement and migration of mature and immature individuals; and,
- An internationally co-ordinated, dedicated anglerfish survey over the wider Northern European area to include waters further east. Currently the Scottish-Irish survey provides a biomass estimate for the whole of VIa, but there is only partial coverage of the North Sea. The survey should be expanded to cover the entire distribution of the stock and this would require the participation of other nations.


### 5.2.1.9 Recommendations for the next benchmark

ICES has previously advised a two-stage approach for management of the anglerfish fishery. The first stage was to substantially improve the quality and quantity of data collected in the fishery while maintaining exploitation at its current level. It has stated that this was expected to take at least five years to establish useable time-series. The second stage would then be to use these data to examine alternative management approaches and harvest control rules. The data collection stage of this process is ongoing and an assessment approach is in preparation. WGCSE 2010 considers that significant progress towards assessment has been made for this stock which is still on track for a benchmark meeting in 2012.

The biological data associated with the anglerfish surveys should be evaluated and compared with existing estimates (e.g. maturity-at-age, growth rates, length distributions, sex ratios and species compositions). There are still uncertainties about the validity of age readings of anglerfish: this will be addressed by an age determination exchange and workshop. Depending on the outcome of this workshop, the catch-atage data should then be evaluated for use in any assessment.

Irrespective of any ageing concerns, the survey estimates have underestimated the younger ages. This is in spite of the recent incorporation of a correction to account for escapes of small fish under the footrope of the survey trawl, which clearly has not accounted for all small fish. Some developments of the latter bias correction are still possible; however, it seems likely that a survey based assessment model could also be developed to determine the absolute abundance of the total population.

A number of recommendations were made at ICES WKAGME for the improvement of the anglerfish surveys. Some of these have been addressed and other will be addressed in the coming year in advance of the Benchmark. These include: improving the survey design in the light of previous estimates of density (allocation of samples to strata); providing estimates for the two species separately so that they may be incorporated separately in any assessment model (for cohort tracking for example); incorporating better procedures for [the few] missing ages; accounting for areas not surveyed in the North Sea using IBTS data; and improving the estimates of footrope escapes.

Finally, it should be stressed that, to date, efforts to extrapolate estimates of abundance into areas that have not been surveyed (southern North Sea and Subarea IIIa) have not proved particularly successful. Additional participation of nations with an interest in this fishery should be encouraged before the next Benchmark. In 2009 only

Scotland and Ireland participated in this survey and in 2010 only Scotland was able to conduct a survey.

### 5.2.2 Anglerfish in Division Ila

The WGNSDS considered the stock structure on a wider European scale in 2004, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division IIa. Anglerfish in IIa is therefore treated in this separate chapter.

## Type of assessment in 2010

No assessment was performed.

## ICES advice applicable to 2009 and 2010

The ICES advice for 2010 (Single Stock Exploitation Boundaries) was as follows, and applies to Subarea VI, Subarea IV, Division IIIa and Division IIa:

ICES advises on the basis of precautionary considerations that the effort in fisheries that catch anglerfish should not be allowed to increase.

### 5.2.2.1 General

## Stock description and management units

The WGNSDS considered the stock structure on a wider European scale in 2004, and found no conclusive evidence to indicate an extension of the stock area northwards to include Division IIa. Anglerfish in IIa is therefore treated in this separate chapter.

## Fishery in 2009

There has been an expansion of the fishery in recent years. This is largely due to a northward expansion of the Norwegian gillnet fishery. Norway is by far the largest exploiter of the IIa fishery accounting for over $95 \%$ of official landings. UK is now the next most important exploiter in this area, with landings of approximately $2.5 \%$ of the total reported to ICES (Table 5.2.10). The coastal gillnetting accounts for 85-90\% of the landings, while $4-6 \%$ is taken as bycatch in different offshore gillnet fisheries (Table 5.2.11).

No TAC is given for Division IIa, Norwegian waters. Catches of anglerfish in Division IIa, EC waters are taken as a part of the TAC for Subarea IV. The Norwegian fishery is regulated through:

- A prohibition against targeting anglerfish with other fishing gear than 360 mm gillnets. A discard ban on anglerfish regardless of size.
- A maximum of $10 \%$ bycatch of anglerfish in the shrimp trawl fishery, maximum $20 \%$ bycatch of anglerfish in the trawl and Danish seine fishery.
- 48 hours maximum soak time in the gillnet fishery
- A maximum of 500 gillnets (each net being 27.5 m ) per vessel.
- A closure of the gillnet fishery from 1 March to 20 May. This closure period was expanded to 20 December to 20 May in the areas north of $\mathrm{N} 65^{\circ}$ in 2008 and this area was expanded southwards to $\mathrm{N} 64^{\circ}$ in 2009.


### 5.2.2.2 Data

## Landings

The official landings for each country are shown in Table 5.2.10. Landings in 2008 as reported to ICES for the total Division IIa were 4447 t , which is 300 t higher than the year before. No information suggests that the official landing figures from Norway give a biased estimate of the actual landings.

## Discards

The absence of a TAC in Norwegian waters probably reduces the incentive to underreport landings. Anecdotal evidence from the industry, observer trips and data from the self-sampling-fleet suggest that a small percentage of the catch (not marketable) is discarded. This happens when the soaking time is too long, mostly due to bad weather. Data are not adequate for estimating discard levels yet.

## Biological

Length distributions are available from the directed gillnetting during the period 1992-2009, but data is lacking 1997-2001 (Figure 5.2.11). The length data indicates a decrease in mean length of $15-20 \mathrm{~cm}$ occurred during the period without length samples. The mean length has increased somewhat during the last five years, but is still below the level seen during the 1990s (Figure 5.2.12). One third of the anglerfish measured during the 1990s were above 100 cm , this proportion was between $1-6 \%$ for the early 2000 s and $14-17 \%$ in 2006-2009. For 2006-2009, some length data from anglerfish caught as bycatch in other fisheries are presented in Figure 5.2.13.

## Surveys

Anglerfish appears in demersal trawl surveys along the Norwegian shelf, but in very low numbers. There has been a change in the surveys, going from single species to multispecies surveys, during recent years. The procedures for data collection on anglerfish have varied and, at present, no time-series from surveys in Division IIa yields reliable information on the abundance of anglerfish.

## Commercial cpue

Reliable effort data are not available from the Norwegian gillnetters due to nonmandatory effort recording. In late 2005, ten gillnetters were included in a selfsampling scheme established along the Norwegian coast within Division IIa. Detailed information about effort and catch is provided through this scheme, and will potentially be valuable in future assessments of anglerfish in this area. The time-series was examined prior to WGCSE 2010, and this revealed some data quality problems which have to be solved before any further analysis.

### 5.2.2.3 Historical stock development

Anglerfish in Division IIa have never been assessed quantitatively and it is not possible to describe the historical stock development.

### 5.2.2.4 Management considerations

The WG notes the apparent changes in size composition in anglerfish caught in the gillnet fishery. If the selectivity in the gillnets has been stable, this could be interpreted as an altering of the size spectrum in the stock. As the information on trends in effort is lacking for the main fishery, it remains unclear whether the increased land-
ings last year might reflect an increased abundance in the area. Time-series on effort and catch by length should be established to facilitate future analytical assessments of this stock. The possibility of establishing a survey, similar to the one being carried out for the Northern Shelf area, should also be considered for Division IIa.

### 5.2.2.5 References

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Table 5.2.1. Anglerfish on the Northern Shelf (IIIa, IV \& VI). Total official landings by area (tonnes).

|  | IIIa | IVa | IVb | IVc | VIa | VIb | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | 140 | 2085 | 575 | 41 | 9221 | 127 | 12189 |
| 1974 | 202 | 2737 | 1171 | 39 | 3217 | 435 | 7801 |
| 1975 | 291 | 2887 | 1864 | 59 | 3122 | 76 | 8299 |
| 1976 | 641 | 3624 | 1252 | 49 | 3383 | 72 | 9021 |
| 1977 | 643 | 3264 | 1278 | 54 | 3457 | 78 | 8774 |
| 1978 | 509 | 3111 | 1260 | 72 | 3117 | 103 | 8172 |
| 1979 | 687 | 2972 | 1578 | 112 | 2745 | 29 | 8123 |
| 1980 | 652 | 3450 | 1374 | 175 | 2634 | 200 | 8485 |
| 1981 | 549 | 2472 | 752 | 132 | 1387 | 331 | 5623 |
| 1982 | 529 | 2214 | 654 | 99 | 3154 | 454 | 7104 |
| 1983 | 506 | 2465 | 1540 | 181 | 3417 | 433 | 8542 |
| 1984 | 568 | 3874 | 1803 | 188 | 3935 | 707 | 11075 |
| 1985 | 578 | 4569 | 1798 | 77 | 4043 | 1013 | 12078 |
| 1986 | 524 | 5594 | 1762 | 47 | 3090 | 1326 | 12343 |
| 1987 | 589 | 7705 | 1768 | 66 | 3955 | 1294 | 15377 |
| 1988 | 347 | 7737 | 2061 | 95 | 6003 | 1730 | 17973 |
| 1989 | 334 | 7868 | 2121 | 86 | 5729 | 313 | 16451 |
| 1990 | 570 | 8387 | 2177 | 34 | 5615 | 822 | 17605 |
| 1991 | 595 | 9235 | 2522 | 26 | 5061 | 923 | 18362 |
| 1992 | 938 | 10209 | 3053 | 39 | 5479 | 1089 | 20807 |
| 1993 | 843 | 12309 | 3144 | 66 | 5553 | 681 | 22596 |
| 1994 | 811 | 14505 | 3445 | 210 | 5273 | 777 | 25021 |
| 1995 | 823 | 17891 | 2627 | 402 | 6354 | 830 | 28927 |
| 1996 | 702 | 25176 | 1847 | 304 | 6408 | 602 | 35039 |
| 1997 | 776 | 23425 | 2172 | 160 | 5330 | 899 | 32762 |
| 1998 | 626 | 16857 | 2088 | 78 | 4506 | 900 | 25055 |
| 1999 | 660 | 13326 | 1517 | 24 | 4284 | 1401 | 21212 |
| 2000 | 602 | 12338 | 1617 | 31 | 3311 | 1074 | 18973 |
| 2001 | 621 | 12861 | 1832 | 21 | 2660 | 1309 | 19304 |
| 2002 | 667 | 11048 | 1244 | 21 | 2280 | 718 | 15978 |
| 2003 | 478 | 8523 | 847 | 20 | 2493 | 643 | 13004 |
| 2004 | 519 | 8987 | 851 | 15 | 2453 | 671 | 13496 |
| 2005 | 458 | 8424 | 688 | 5 | 3019 | 958 | 13552 |
| 2006 | 423 | 10338 | 685 | 3 | 2785 | 916 | 15150 |
| 2007 | 433 | 10632 | 749 | 4 | 3352 | 1260 | 16430 |
| 2008 | 486 | 11038 | 769 | 5 | 3373 | 1630 | 17300 |
| 2009 | 548 | 10286 | 752 | 8 | 3178 | 1767 | 16539 |
| Min | 140 | 2085 | 575 | 3 | 1387 | 29 | 5623 |
| Max | 938 | 25176 | 3445 | 402 | 9221 | 1767 | 35039 |
| Average | 564 | 8768 | 1601 | 82 | 4010 | 773 | 15798 |

## Table 5.2.2. Anglerfish in Subarea VI. Nominal landings ( $\mathbf{t}$ ) as officially reported to ICES.

Anglerfish in Division VIa (West of Scotland)

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 2 | 9 | 6 | 5 | - | 5 | 2 | - | - | + | + | - | + | - | - | - | - | - |
| Denmark | 1 | 3 | 4 | 5 | 10 | 4 | 1 | 2 | 1 | + | + | . | + | + | - | - | - | - | + |
| Faroe Is. | - | - | - | - | - | - | - | - | - | - | - | - | - | 2 | 2 | 3 | 2 | 1 |  |
| France | 1,910 | 2,308 | 2,467 | 2,382 | 2,648 | 2,899 | 2,058 | 1,634 | 1,814 | 1,132 | 943 | 739 | 1,212 | 1,191 | 1,392 | 1,314 | 1763 | 1746 | 1928 |
| Germany | 1 | 2 | 60 | 67 | 77 | 35 | 72 | 137 | 50 | 39 | 11 | 3 | 27 | 39 | 39 | 1 | - | 54 | 79 |
| Ireland | 250 | 403 | 428 | 303 | 720 | 717 | 625 | 749 | 617 | 515 | 475 | 304 | 322 | 219 | 356 | 392 | 470 | 295 | 328 |
| Netherlands | - | - | - | - | - | - | 27 | 1 | - | - | - | - | - | - | - | - | - | - |  |
| Norway | 6 | 14 | 8 | 6 | 4 | 4 | 1 | 3 | 1 | 3 | 2 | 1 | + | + | 1 | 1 | 1 | 2 | - |
| Spain | 7 | 11 | 8 | 1 | 37 | 33 | 63 | 86 | 53 | 82 | 70 | 101 | 196 | 110 | 82 | 76 | 3 | 174 |  |
| UK(E,W\&NI) | 270 | 351 | 223 | 370 | 320 | 201 | 156 | 119 | 60 | 44 | 40 | 32 | 31 | 30 | 20 | 24 | 42 | 5 |  |
| UK(Scot.) | 2,613 | 2,385 | 2,346 | 2,133 | 2533 | 2,515 | 2,322 | 1,773 | 1,688 | 1,496 | 1,119 | 1,100 | 705 | 862 | 1,127 | 974 | 1,071 | 1096 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 876 |
| Total | 5,061 | 5,479 | 5,553 | 5,273 | 6,354 | 6,408 | 5,330 | 4,506 | 4,284 | 3,311 | 2,660 | 2,280 | 2,493 | 2,453 | 3,019 | 2,785 | 3,352 | 3,373 | 3,211 |
| Unallocated | 296 | 2,638 | 3,816 | 2,766 | 5,112 | 11,148 | 7,506 | 5,234 | 3,799 | 3,114 | 2,068 | 1,882 | 985 | 1,938 |  |  |  |  |  |
| As used by WG | 5,357 | 8,117 | 9,369 | 8,039 | 11,466 | 17,556 | 12,836 | 9,740 | 8,083 | 6,425 | 4,728 | 4,162 | 3,478 | 4,391 |  |  |  |  |  |

*Preliminary.

Table 5.2.2 contd. Anglerfish in Subarea VI. Nominal landings ( $\mathbf{t}$ ) as officially reported to ICES.
Anglerfish in Division VIb (Rockall)

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estonia | - | - | - | - | - | - | - | - | - | - | - | - | - | + | - | - | - |  |  |
| Faroe Is. | - | 2 | - | - | - | 15 | 4 | 2 | 2 | - | 1 | - | - | - | - | - | - | 1 |  |
| France | - | - | 29 | - | - | - | 1 | 1 | ... 1 | 48 | 192 | 43 | 191 | 175 | 293 | 224 | 327 | 327 | 361 |
| Germany | - | - | 103 | 73 | 83 | 78 | 177 | 132 | 144 | 119 | 67 | 35 | 64 | 66 | 77 | 72 | 222 | 0 | 132 |
| Ireland | 272 | 417 | 96 | 135 | 133 | 90 | 139 | 130 | 75 | 81 | 134 | 51 | 26 | 13 | 35 | 53 | 70 | 76 | 91 |
| Norway | 18 | 10 | 17 | 24 | 14 | 11 | 4 | 6 | 5 | 11 | 5 | 3 | 6 | 5 | 4 | 6 | 7 | 5 | 9 |
| Portugal | - | - | - | - | - | - | - | + | 429 | 20 | 18 | 8 | 4 | 19 | 63 | - | - | - |  |
| Russia | - | - | - | - | - | - | - | - | - | - | 1 | - | - | 2 | 4 | 1 | 1 | 35 |  |
| Spain | 333 | 263 | 178 | 214 | 296 | 196 | 171 | 252 | 291 | 149 | 327 | 128 | 59 | 43 | 34 | 36 | 12 | 85 |  |
| UK(E,W\&NI) | 99 | 173 | 76 | 50 | 105 | 144 | 247 | 188 | 111 | 272 | 197 | 133 | 133 | 54 | 93 | 46 | 146 | 5 |  |
| UK(Scot) | 201 | 224 | 182 | 281 | 199 | 68 | 156 | 189 | 344 | 374 | 367 | 317 | 160 | 294 | 355 | 478 | 475 | 1096 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1189 |
| Total | 923 | 1089 | 681 | 777 | 830 | 602 | 899 | 900 | 1401 | 1074 | 1309 | 718 | 643 | 671 | 958 | 916 | 1260 | 1630 | 1767 |
| Unallocated |  |  |  |  |  |  |  |  | -9 | 17 | -178 | -47 | 145 | 121 |  |  |  |  |  |
| As used by WG | 923 | 1,089 | 681 | 777 | 830 | 602 | 899 | 900 | 1392 | 1091 | 1131 | 671 | 788 | 792 |  |  |  |  |  |

*Preliminary.

Total Anglerfish in Sub-area VI (West of Scotland and Rockall)

| Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total official | 5,984 | 6,568 | 6,234 | 6,050 | 7,184 | 7,010 | 6,229 | 5,406 | 5,685 | 4,385 | 3,969 | 2,998 | 3,136 | 3,124 | 3,977 | 3,701 | 5,003 | 4,945 |
| Total <br> ICES | 6,280 | 9,206 | 10,050 | 8,816 | 12,296 | 18,158 | 13,735 | 10,640 | 9,475 | 7,516 | 5,859 | 4,833 | 4,266 | 5,183 |  |  |  |  |

*Preliminary.
Table 5.2.3. Nominal landings (t) of Anglerfish in the North Sea, as officially reported to ICES.
Northern North Sea (IVa)

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 2 | 9 | 3 | 3 | 2 | 8 | 4 | 1 | 5 | 12 | - | 8 | 1 | - | - | - | - | - |  |
| Denmark | 1,245 | 1265 | 946 | 1,157 | 732 | 1,239 | 1,155 | 1,024 | 1,128 | 1,087 | 1,289 | 1,308 | 1,523 | 1,538 | 1,379 | 1,311 | 961 | 1,071 | 1,356 |
| Faroes | 1 | - | 10 | 18 | 20 | - | 15 | 10 | 6 | . | 2 | + | 3 | 11 | 22 | 2 | + | - |  |
| France | 124 | 151 | 69 | 28 | 18 | 7 | 7 | 3* | 181* | 8 | 9 | 8 | 8 | 8 | 4 | 7 | 13 | 13 |  |
| Germany | 71 | 68 | 100 | 84 | 613 | 292 | 601 | 873 | 454 | 182 | 95 | 95 | 65 | 20 | 84 | 173 | 186 | 344 | 216 |
| Netherlands | 23 | 44 | 78 | 38 | 13 | 25 | 12 | - | 15 | 12 | 3 | 8 | 9 | 38 | 13 | 14 | 14 | 12 | 6 |
| Norway | 587 | 635 | 1,224 | 1,318 | 657 | 821 | 672 | 954 | 1,219 | 1,182 | 1,212 | 928 | 769 | 999 | 880 | 1,005 | 831 | 860 | 859 |
| Sweden | 14 | 7 | 7 | 7 | 2 | 1 | 2 | 8 | 8 | 78 | 44 | 56 | 8 | 6 | 5 | 5 | 20 | 67 | 21 |
| UK(E, | 129 | 143 | 160 | 169 | 176 | 439 | 2,174 | 668 | 781 | 218 | 183 | 98 | 104 | 83 | 34 | 99 | 303 | 13 |  |
| UK | 7,039 | 7,887 | 9,712 | 11,683 | 15,658 | 22,344 | 18,783 | 13,319 | 9,710 | 9,559 | 10,024 | 8,539 | 6,033 | 6,284 | 6,003 | 7,722 | 8304 | 8,658 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 7,828 |
| Total | 9,235 | 10,209 | 12,309 | 14,505 | 17,891 | 25,176 | 23,425 | 16,857 | 13,326 | 12,338 | 12,861 | 11,048 | 8,523 | 8,987 | 8,424 | 10,338 | 10,632 | 11,038 | 10,286 |

* Preliminary.

Table 5.2.3 continued. Nominal landings ( t ) of Anglerfish in the North Sea as officially reported to ICES.
Central North Sea (IVb)

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 357 | 538 | 558 | 713 | 579 | 287 | 336 | 371 | 270 | 449 | 579 | 435 | 180 | 260 | 207 | 138 | 179 | 181 | 133 |
| Denmark | 345 | 421 | 347 | 350 | 295 | 225 | 334 | 432 | 368 | 260 | 251 | 255 | 191 | 274 | 237 | 276 | 173 | 237 | 337 |
| Faroes | - | - | 2 | - | - | - | - | - | - | - | - | 10 | - | - | - | - | - | - |  |
| France | - | 1 | - | 2 | - | - | - | -* | $\ldots 1^{*}$ | - | - | - | - | + | - | - | - | - |  |
| Germany | 4 | 2 | 13 | 15 | 10 | 9 | 18 | 19 | 9 | 14 | 9 | 17 | 11 | 11 | 9 | 14 | 12 | 22 | 17 |
| Ireland |  |  |  |  |  |  |  |  |  |  |  |  | 1 | - | - | - | - | - |  |
| Netherlands | 285 | 356 | 467 | 510 | 335 | 159 | 237 | 223 | 141 | 141 | 123 | 62 | 42 | 25 | 31 | 33 | 61 | 58 | 47 |
| Norway | 17 | 4 | 3 | 11 | 15 | 29 | 6 | 13 | 17 | 9 | 15 | 10 | 12 | 22 | 16 | 14 | 24 | 15 | 21 |
| Sweden | - | - | - | 3 | 2 | 1 | 3 | 3 | 4 | 3 | 2 | 9 | 2 | 1 | 4 | 4 | 6 | 9 | 6 |
| UK(E, W\&NI) | 669 | 998 | 1,285 | 1,277 | 919 | 662 | 664 | 603 | 364 | 423 | 475 | 236 | 167 | 120 | 96 | 108 | 122 | 105 |  |
| UK (Scotland) | 845 | 733 | 469 | 564 | 472 | 475 | 574 | 424 | 344 | 318 | 378 | 210 | 241 | 138 | 88 | 98 | 172 | 142 |  |
| UK (total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 191 |
| Total | 2,522 | 3,053 | 3,144 | 3,445 | 2,627 | 1,847 | 2,172 | 2,088 | 1,517 | 1,617 | 1,832 | 1,244 | 847 | 851 | 688 | 685 | 749 | 769 | 752 |

* Preliminary

Table 5.2.3 continued. Nominal landings ( t ) of Anglerfish in the North Sea as officially reported to ICES.

Southern North Sea (IVc)

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 13 | 12 | 34 | 37 | 26 | 28 | 17 | 17 | 11 | 15 | 15 | 16 | 9 | 5 | 4 | 3 | 3 | 4 | 6 |
| Denmark | 2 | + | - | + | + | + | + | + | + | + | + | + | + | + | + | - | - |  |  |
| France | - | - | - | - | - | - | - | 10 | - | + | - | + | - | - | - | - | - | + |  |
| Germany | - | - | - | - | - | - | - | - | - | + | - | + | + | - | - | - | - | - | + |
| Netherlands | 5 | 10 | 14 | 20 | 15 | 17 | 11 | 15 | 10 | 15 | 6 | 5 | 1 | - | 1 | - | 1 | 1 |  |
| Norway | - | - | - | - | + | - | - | - | + | - | + | - | - | - | - | - | - | - | 1 |
| UK(E\&W\&NI) | 6 | 17 | 18 | 136 | 361 | 256 | 131 | 36 | 3 | 1 | - | - | 10 | 3 | - | - | - | ... |  |
| UK (Scotland) | - | - | - | 17 | - | 3 | 1 | + | + | + | - | - | - | 7 | - | - | - | ... |  |
| UK (Total) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | + | 1 |
| Total | 26 | 39 | 66 | 210 | 402 | 304 | 160 | 78 | 24 | 31 | 21 | 21 | 20 | 15 | 5 | 3 | 4 | 5 | 8 |

* Preliminary.

Total North Sea

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total | 11,783 | 13,301 | 15,519 | 18,162 | 20,920 | 27,327 | 25,757 | 19,023 | 14,867 | 13,986 | 14,714 | 12,313 | 9,390 | 9,853 | 9,117 | 11,026 | 11,385 | 11,812 | 11,046 |
| WG estimate | 10,566 | 11,728 | 13,078 | 15,432 | 15,794 | 16,240 | 18,217 | 14,027 | 11,719 | 11,564 | 12,677 | 10,334 | 8,273 | 9,027 |  |  |  |  |  |
| Unallocated | -1,217 | -1,573 | -2,441 | $-2,730$ | -5,126 | $11,087$ | -7,540 | -4,996 | -3,148 | -2,422 | -2,037 | -1,979 | $1,117$ | -826 |  |  |  |  |  |

* Preliminary.

Table 5.2.4. Nominal landings (t) of Anglerfish in Division IIIa, 1991-2009, as officially reported to ICES.

|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 15 | 48 | 34 | 21 | 35 | - | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Denmark | 493 | 658 | 565 | 459 | 312 | 367 | 550 | 415 | 362 | 377 | 375 | 369 | 215 | 311 | 274 | 227 | 255 | 287 | 372 |
| Germany | - | - | 1 | - | - | 1 | 1 | 1 | 2 | 1 | - | 1 | - | 1 | 1 | 2 | 1 | 1 | 1 |
| Netherlands |  |  |  |  |  |  | - | - | - | - | - | . | 3 | 4 | 4 | 3 | 1 | 3 |  |
| Norway | 64 | 170 | 154 | 263 | 440 | 309 | 186 | 177 | 260 | 197 | 200 | 242 | 189 | 130 | 100 | 137 | 132 | 144 | 134 |
| Sweden | 23 | 62 | 89 | 68 | 36 | 25 | 39 | 33 | 36 | 27 | 46 | 55 | 71 | 73 | 79 | 54 | 44 | 51 | 41 |
| Total | 595 | 938 | 843 | 811 | 823 | 702 | 776 | 626 | 660 | 602 | 621 | 667 | 478 | 519 | 458 | 423 | 433 | 486 | 548 |

*Preliminary.

Tables 5.2.5. Total Danish Anglerfish landings (tonnes) and effort (days fishing) by fishery.
A. Landings by fishery (from logbook data)

B. Effort by fishery (from logbook data)

| Year |  |  |  |  | Total Danish effort in IV (days) | North Sea |  |  |  |  | Total Danish effort in IIIA (days) |  | IIIa | IIIa \& IV |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Other gear | Beam trawls | dem <br> trawl | Neph <br> trawl | ind trawl | Shrimp trawl | total | Other gear | Beam trawls | dem <br> trawl | Neph <br> trawl | ind trawl | Shrimp trawl | total | total |
| 2000 | 695 | 787 | 6297 | 285 | 808 | 1102 | 9974 | 316 | 410 | 962 | 2173 | 5 | 227 | 4092 | 14066 |
| 2001 | 780 | 250 | 8164 | 182 | 1039 | 1137 | 11552 | 315 | 267 | 775 | 2916 | 31 | 219 | 4522 | 16074 |
| 2002 | 676 | 537 | 7415 | 741 | 1155 | 1025 | 11548 | 297 | 356 | 1054 | 2570 | 18 | 210 | 4505 | 16053 |
| 2003 | 309 | 445 | 7917 | 711 | 528 | 810 | 10720 | 174 | 62 | 328 | 1983 | 7 | 188 | 2742 | 13462 |
| 2004 | 522 | 419 | 6212 | 448 | 517 | 606 | 8725 | 309 | 165 | 211 | 2638 | 3 | 135 | 3462 | 12186 |
| 2005 | 166 | 401 | 6075 | 443 | 240 | 263 | 7589 | 141 | 92 | 517 | 1991 | 3 | 154 | 2898 | 10487 |
| 2006 | 174 | 96 | 5912 | 543 | 125 | 154 | 7004 | 99 | 43 | 539 | 1403 | 2 | 52 | 2139 | 9143 |
| 2007 | 108 | 191 | 3805 | 361 | 106 | 36 | 4607 | 117 | 139 | 744 | 1244 | 0 | 181 | 2424 | 7031 |
| 2008 | 189 | 191 | 3978 | 469 | 38 | 104 | 4968 | 185 | 51 | 690 | 2031 | 1 | 397 | 3356 | 8325 |
| 2009 | 414 | 215 | 4823 | 443 | 14 | 166 | 6076 | 178 | 45 | 821 | 3027 | 15 | 582 | 4668 | 10744 |

## Table 5.2.6. Anglerfish in IV and IIIa. Norwegian landings (tonnes) by fishery in 2005-2008 and preliminary data from 2009.

| Fleet | 2005 Div IIIA | 2005 Div IVA | 2006 Div IIIA | 2006 Div IVA | 2007 Div IIIA | 2007 Div Iva | 2008 Div IIIA | 2008 DIv IVA | 009 DIv IIIA | 2009 Div Iva |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Coastal gillnetting | 61 | 526 | 103 | 696 | 87 | 574 | 97 | 554 | 90 | 481 |
| Offshore gillnetting | 1 | 16 | + | 19 | + | 32 | + | 24 | + | 21 |
| Coastal shrimp trawling | 22 | 50 | 25 | 46 | 26 | 36 | 27 | 35 | 30 | 29 |
| Offshore dem trawling | 5 | 102 | + | 142 | 8 | 154 | 12 | 206 | 6 | 265 |
| Offshore shrimp trawling | 3 | 68 | 5 | 66 | 8 | 39 | 7 | 32 | 6 | 40 |
| Other gears | 7 | 119 | 3 | 36 | 3 | 24 | + | 24 | 2 | 23 |
| Total | 100 | 880 | 137 | 1,005 | 132 | 860 | 144 | 875 | 134 | 859 |

Table 5.2.7. Anglerfish in Subarea VI. Landings, effort and lpue from the Irish OTB fleet.

| Year | Hours (VIa) | Kw.Days (VIa) | Hours VIb) | kw.Days (VIb) | Landings (Vla) | Landings (VIb) | LPUE <br> (VIa_Hours) | LPUE (VIa kw.days) | LPUE <br> (VIb_Hours) | LPUE (VIb kw.days) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 56863 | 1408312 | 9029 | 599053 | 655 | 114 | 11.52 | 0.47 | 12.63 | 0.019 |
| 1996 | 60960 | 1388902 | 7219 | 469212 | 624 | 74 | 10.24 | 0.45 | 10.25 | 0.022 |
| 1997 | 63159 | 1462368 | 7169 | 377836 | 587 | 93 | 9.29 | 0.40 | 12.97 | 0.025 |
| 1998 | 57398 | 1343782 | 7337 | 403310 | 558 | 99 | 9.72 | 0.42 | 13.49 | 0.024 |
| 1999 | 54075 | 1348480 | 8680 | 437920 | 449 | 64 | 8.30 | 0.33 | 7.37 | 0.019 |
| 2000 | 52847 | 1325585 | 9883 | 613229 | 410 | 62 | 7.76 | 0.31 | 6.27 | 0.013 |
| 2001 | 47224 | 1320179 | 7232 | 593467 | 315 | 93 | 6.67 | 0.24 | 12.86 | 0.011 |
| 2002 | 35016 | 1007965 | 2626 | 217918 | 276 | 41 | 7.88 | 0.27 | 15.61 | 0.036 |
| 2003 | 39211 | 1536279 | 4543 | 478464 | 314 | 26 | 8.01 | 0.20 | 5.72 | 0.017 |
| 2004 | 35217 | 1279049 | 2234 | 205349 | 210 | 13 | 5.96 | 0.16 | 5.82 | 0.029 |
| 2005 | 30748 | 1075974 | 3844 | 216991 | 351 | 35 | 11.42 | 0.33 | 9.11 | 0.053 |
| 2006 | 28014 | 1031169 | 5903 | 464965 | 386 | 53 | 13.78 | 0.37 | 8.98 | 0.030 |
| 2007 | 25373 | 911973 | 6589 | 548392 | 467 | 69 | 18.41 | 0.51 | 10.47 | 0.034 |
| 2008 | 17327 | 630615 | 9740 | n/a | 295 | 78 | 17.03 | 0.47 | 8.01 | n/a |
| 2009 | 17107.5 |  | 4354 |  | 331.632 | 91 | 19.39 |  | 20.90 |  |

Landings in tonnes
Lpue estimates on ' 000 hours fished or ' 000 kw.days

Table 5.2.8. Danish lpue ( $\mathrm{Kg} /$ day) for anglerfish. Official logbook records and for shrimp trawl adjusted for increasing fishing power (technological creep).

|  |  | North sea (IV) \& Skagerrak (IIIa) |  | North sea (IV) \& Skagerrak (IIIa) |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Lpue, nominal (unadjusted) |  | Relative lpue, nominal (unadjusted) |  | Relative lpue, adjusted for increasing fishing power |
|  | Dem. Trawl + <br> Neph trawl | Shrimp trawl | Dem. Trawl + <br> Neph trawl | Shrimp trawl | Shrimp trawl |
| 1996 | 176.1 | 91.0 | 1.03 | 1.49 | 2.82 |
| $1997$ | 170.6 | 61.0 | 1.00 | 1.00 | 1.00 |
| 1998 | 176.6 | 59.4 | 1.03 | 0.97 | 1.27 |
| 1999 | 163.5 | 62.3 | 0.96 | 1.02 | 0.95 |
| 2000 | 132.2 | 49.0 | 0.77 | 0.80 | 0.73 |
| 2001 | 131.7 | 45.3 | 0.77 | 0.74 | 0.67 |
| 2002 | 135.3 | 54.5 | 0.79 | 0.89 | 0.62 |
| 2003 | 156.4 | 49.4 | 0.92 | 0.81 | 0.47 |
| 2004 | 194.3 | 75.4 | 1.14 | 1.24 | 0.54 |
| 2005 | 188.3 | 55.4 | 1.10 | 0.91 | 0.24 |
| 2006 | 205.6 | 54.4 | 1.20 | 0.89 | 0.09 |
| 2007 | 208.1 | 45.0 | 1.22 | 0.74 | 0.09 |
| 2008 | 202.5 | 42.6 | 1.19 | 0.70 | 0.20 |
| 2009 | 211.5 | 39.7 | 1.24 | 0.65 | - |

Table 5.2.9. Abundance (millions of individuals) and biomass (thousands of tonnes) estimates from the 2005-2009 Northern shelf anglerfish surveys by ICES area and division.


Table 5.2.10. Nominal catch ( $\mathbf{t}$ ) of Anglerfish in Division IIa, 1993-2009, as officially reported to ICES.

|  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | + | + | + | + | + | + | + | + | 2 | + | - | 1 | - | - |
| Faroes | + | + | + | + | + | + | + | - | 1 | 1 | 2 | 5 | 11 | 4 |
| France | - | - | - | - | - | - | + | - | - | - | - | - | - | 1 |
| Germany | 2 | 3 | 1 | 4 | 20 | 53 | 4 | 17 | 65 | 59 | 55 | 70 | 55 | - |
| Norway | 3,044 | 1,026 | 526 | 893 | 576 | 1,488 | 1,731 | 2,952 | 3,552 | 2,000 | 2,404 | 2,906 | 2,649 | 4,253 |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - |
| Sweden | - | - | - | + | + | + | + | + | + | - | - | - | - | - |
| UK (total) | 1 | 2 | 74 | 15 | 5 | 7 | 6 | 30 | 2 | 10 | 15 | 18 | 19 | 86 |
| Total | 3,047 | 1,031 | 601 | 912 | 601 | 1,548 | 1,741 | 2,999 | 3,622 | 2,070 | 2,476 | 2,999 | 2,672 | 4,341 |

Table 5.2.10. continued.

|  | 2007 | 2008 | 2009* |
| :--- | :---: | :---: | :---: |
| Denmark | - | - | + |
| Faroes | 7 | 4 |  |
| France | - | - |  |
| Germany | - | - | - |
| Norway | 4,455 | 3,999 | 4,289 |
| Portugal | - | 2 | 6 |
| Russia | - | - | - |
| Sweden | - | - | - |
| UK (total) | 115 | 138 | 152 |
| Total | 4,577 | 4,143 | 4,447 |

*Preliminary.

Table 5.2.11. Anglerfish in IIa. Norwegian landings (tonnes) by fishery in 2005-2008 and preliminary data for 2009.

| FLEET | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Coastal <br> gillnetting | 2,301 | 3,723 | 4,039 | 3,574 | 3,934 |
| Offshore <br> gillnetting | 115 | 261 | 204 | 240 | 172 |
| Offshore dem <br> trawling | 77 | 71 | 52 | 26 | 28 |
| Coastal <br> Danish seine | 54 | 54 | 63 | 75 | 68 |
| Other gears | 102 | 144 | 98 | 84 | 87 |
| Total | 2,649 | 4,253 | 456 | 3,999 | 4,289 |



Figure 5.2.1. Northern Shelf anglerfish. Officially reported landings by ICES area.



Figure 5.2.2. Danish landings of Anglerfish by fishery in the North Sea (top) and Division IIIa (bottom).


Figure 5.2.3. Anglerfish in Division IVa. Norwegian landings by fleet from 2003-2009.


Figure 5.2.4. Map of the European Northern Shelf showing the distribution of reported landings of anglerfish for 2009 from Scotland, Ireland, Denmark, France and Norway. The red circles are centred on each ICES rectangle and the area of each circle is proportional to the landings in tonnes (according to the legend). The data have been corrected according to certain assumptions about area misreporting (see Stock Annex).


Figure 5.2.5. Lpue for the Irish otter trawl fleet with effort in hours fished for a) Division VIa, and b) Division VIb.


Figure 5.2.6. Anglerfish in the North Sea \& Division IIIa. Danish lpue by demersal trawl and shrimp trawl, relative to 1997. Based on nominal logbook records as well as development in gear and engine power (shrimp trawl).


Figure 5.2.7. Map of the northern continental shelf around Scotland showing the number density of anglerfish during the 2009 surveys. Each circle is centred on the sample location and circle size is proportional to the number density in $\mathbf{n} / \mathbf{k m}^{2}$ according to the legend (top left). Blue circles represent trawl based densities based on Scottish surveys; green symbols Irish surveys. Trawl densities in this figure account for herding but not footrope escapes. The red lines separate the ICES subareas indicated by roman numerals: IV (east) and VI (west).


Figure 5.2.8. Map of the northern continental shelf around Scotland showing the weight density of anglerfish during the 2009 anglerfish survey. Each circle is centred on the sample location and circle size is proportional to weight density in $\mathrm{kg} / \mathrm{km}^{2}$ according to the legend. Blue circles represent trawl based densities based on Scottish surveys; green circles Irish surveys. Trawl densities in this figure account for herding but not footrope escapes. The red lines separate the ICES subareas indicated by roman numerals: IV (east) and VI (west).


Figure 5.2.9. Estimates of total abundance-at-age for each of the anglerfish surveys 2005-2009. Red bars indicate estimates prior to correction for footrope escapes; blues bars include the latter correction; green bars indicate an additional correction for the unsurveyed part of ICES Division VIa based on data when the area was surveyed by the Irish. Error bars are $\mathbf{9 5 \%}$ confidence intervals.


Figure 5.2.10. Estimates of total abundance (left) and biomass (right) of anglerfish for the Northern shelf (black filled circles), with confidence intervals derived from variance estimates of the Scottish surveys. Estimates are also provided for ICES Subarea IV (red filled squares), Division VIa (blue open circles) and Division VIb (green filled triangles). Confidence limits for 2005 biomass are provisional.


Figure 5.2.11. Anglerfish in IIa. Length distributions for anglerfish caught in the directed coastal gillnetting in Division IIa during 1992-2009. Note that data are lacking for 1997-2001.


Figure 5.2.12. Anglerfish in IIa. Mean lengths for anglerfish caught in the directed coastal gillnetting in Division IIa during 1992-2009, dotted lines represents $\pm 2$ SE of the mean. Note that data are lacking for 1997-2001.


Figure 5.2.13. Anglerfish in IIa. Length distribution for anglerfish caught as bycatch by other gears (offshore gillnetting and longlining) in Division IIa in 2006-2009.

### 5.3 Megrim in Division IV and VI

Type of assessment in 2010
ICES has not conducted an analytical assessment of this stock since 1999. Megrim continues to be a monitored stock and a benchmark analysis will be required before an assessment can be presented. Based on the recommendation of WGNSDS (2008), WGCSE now also considers megrim in IVa and IIa.

## ICES advice applicable to 2009

The new landings, cpue, and survey data available for this stock do not change the perception of the stock and do not give reason to change the advice from 2007. The advice on this stock for the fishery in 2009 is therefore the same as the advice given in 2007 for the 2008 fishery: Catches should be based on the recent average (2004 2006), about 1400 t. This includes landings in Division VIa and VIb and unallocated landings in Subarea IV.

## ICES advice applicable to 2010

ICES advises on the basis of exploitation boundaries in relation to precautionary considerations that the effort in fisheries that catch megrim should not be allowed to increase.

### 5.3.1 General

## Stock description and management units

Megrim stock structure is uncertain and historically the Working Group has considered megrim populations in VIa and VIb as separate stocks. The review group questioned the basis for this in 2004. Data collected during an EC study contract (98/096) on the 'Distribution and biology of anglerfish and megrim in the waters to the West of Scotland' showed significantly different growth parameters and significant population structure difference between megrim sampled in VIa and VIb (Anon, 2001). Spawning fish occur in both areas but whether these populations are reproductively isolated is not clear. As noted by WGNSDS (2008), megrim in IVa has historically not been considered by ICES and WGNSDS (2008). Since 2009 data from IV and IIa are now included in this report and work is underway to collect international catch and weight-at-age data for IV as well as VI.

| Species: Megrims <br> Lepidorhombus spp. |  | Zone: | EU waters of IIa and IV (LEZZ/2AC4-C) |
| :---: | :---: | :---: | :---: |
| Belgium | 5 |  |  |
| Denmark | 5 |  |  |
| Germany | 5 |  |  |
| France | 29 |  |  |
| The Netherlands | 23 |  |  |
| United Kingdom | 1690 |  |  |
| EU | 1757 |  |  |
| TAC | 1757 |  | Precautionary TAC |
| Species: Megrims <br> Lepidorhombus spp. |  | Zone: | VI ; EU and international waters of Vb ; international waters of XII and XIV (LELZ/561 214) |
| Spain | 350 |  |  |
| France | 1364 |  |  |
| Ireland | 399 |  |  |
| United Kingdom | 966 |  |  |
| EU | 3079 |  |  |
| TAC | 3079 |  | Precautionary TAC |

## Fishery in 2009

The introduction of the Cod Long-Term Management Plan (EC Regulation 1342/2008) and additional emergency measures applicable to VIa in 2009 (EC Regulation 43/2009, annex III 6) has impacted on the amount of effort deployed and increased the gear selectivity pattern of the main otter trawl fleets. Figure 5.3.1 shows the effort pattern for the main fleets catching megrim in VIa. Additionally, EC regulation 43/2009 has effectively prohibited the use of mesh sizes $<120 \mathrm{~mm}$ for vessels targeting fish, which had been used particularly by the Irish fleet up to that point, the resultant rapid decline in effort for this category can be seen in Figure 5.3.1. Effort associated with the French fleet has continued to decline while the decline in both the Irish and Scottish TR1 fleets ( 120 mm mesh) appears to have stabilized. Note that 2009 data is only available for the Irish fleets. The increase in mesh size (from 100 to 120 mm ) has also impacted on the retention length of megrim, increasing L50 from 28 cm to 42 cm , an increase of almost $50 \%$ (Figure 5.3.2).

Fishing effort in IV for the main Scottish otter fleet (TR1) have stabilized since the large effort reductions observed in previous years, effort levels associated with this mesh band have fallen by $64 \%$ since 2000. Following the increases in Irish effort in subdivision VIb from 2004-2008, effort in 2009 has declined significantly (Figure 5.3.3). There is anecdotal information from the Scottish industry that since the introduction of the Conservation Credits Scheme in Area IV, those vessels have responded with increasing focus on anglerfish and megrim in both IVa and VIa.

Based on landings data presented to the Working Group, only $53 \%$ of the overall TAC for VI, EC waters of Vb and international waters of XII and XIV was used. It should be noted that no landings data were made available to the Working Group by

Spain or France therefore the uptake during 2009 will be higher, while historically, France only utilizes $\sim 10 \%$ of its available quota, Spanish uptake has been $\sim 80 \%$.

2009 TAC for VI, EC waters of Vb and international waters of XII and XIV.

|  | TAC | WG Landings | \% TAC uptake ${ }^{1}$ |
| :--- | :---: | :---: | :---: |
| Spain | 318 | nr |  |
| France | 1240 | nr |  |
| Ireland | 363 |  | $236^{2}$ |
| United Kingdom | 878 |  | 1131 |
| EC Total | 2799 |  | 1380 |

```
*nr - not reported to the Working Group
1 - post regulation quota swaps have not been taken into account
2 Provisional figures
```

The uptake of the TAC for ICES Division IV and IIa was 99\%. Landings data was only received from the UK, which holds $93 \%$ of the TAC.

2009 TAC for EC IV and IIa.

|  | TAC | WG landings | \% TAC uptake ${ }^{1}$ |
| :--- | :---: | :---: | :---: |
| Belgium | 5 | 1.6 | $32 \%$ |
| Denmark | 4 | nr |  |
| Germany | 4 |  | 4 |
| France | 26 | nr |  |
| Netherlands | 21 |  | $100 \%$ |
| UK | 1537 | 2 | $10 \%$ |
| EC | 1597 | 1486 | $96 \%$ |

${ }^{1}$ - post regulation quota swaps have not been taken into account

### 5.3.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1.
It has not been possible to construct full international catch numbers-at-age for the past few years. Data from 2005 to 2007 are required from the UK (VI and IV), France (IV) and Spain (VI). Catch numbers and weights-at-age from 1993 to present have been presented by Ireland (with the exception of 2007 due to lack of market access). Intersessional work is currently underway to construct a full international dataset, with the aim of benchmarking VI and IV megrim in 2010 (see Section 5.3.6).

## Landings

Official landings data for each country together with Working Group best estimates of landings from VIa and VIb and are shown in Table 5.3.1 and landings from IV in Table 5.3.2. The distributions of landings by statistical rectangle from 2007 to 2009 are shown in Figure 5.3.4. The WG best estimates of landings are those supplied by stock coordinators of the various countries and differ from the official statistics in some years. These were supplied for VIa by Ireland and Scotland in 2009. Due to national database problems, France were unable to provide 2009 landings data. Landings have increased in recent years and are more in line with historical trends.

Catches of megrim comprise two species, Lepidorhombus whiffiagonis and L. boscii. Information available to the Working Group indicates that L. boscii, are a negligible proportion of the Scottish and Irish megrim catch (Kunzlik et al., 1995; Anon, 2001). It is not clear to the WG whether landings of other countries are accurately partitioned by megrim species. Megrim are caught in association with anglerfish by some fleets and are area-misreported along with anglerfish. As with anglerfish, the reported $\mathrm{Su}-$ barea VI landings have been adjusted to the Working Groups estimate of catch by including landings declared from Subarea IV in the ICES statistical rectangles immediately east of the 4 degree W line (see anglerfish Annex 5.2 for a detailed methodology). Area-misreporting peaked in 1996 and 1997 when around $50 \%$ of the estimated Working Group landings for Division VIa were area-misreported. This year this correction process has not been conducted. There are indications that more recently the process has reversed. Laurenson and MacDonald (2008) note that in more recent years that megrim TAC in the North Sea has become more restrictive and anecdotal evidence suggest that megrim catches from IV are misreported as coming from Subdivision VI. Therefore, because of conflicting information on the potential direction of area-misreporting, megrim landings at a statistical rectangle level has not been adjusted.

## Discards

Discard data were only made available by Ireland. Discard data from the otter trawl fleet were available for VIa. A mean discard rate of $6 \%$ by weight and $22 \%$ by number is observed, although this is based on data from a limited number of trips (three) and only 44 individual hauls. Laurenson and MacDonlad (2008) note that while discarding of megrim below minimum landing size is low ( $<1 \%$ ), discarding of legal sized fish was much higher at $22 \%$ over the six observed trips. This is attributed to low market price for small grades and bruised fish, resulting in high grading of catches on length/quality reasons to maximise the value of a restrictive quota.

## Surveys

In 2005, Scotland initiated a new industry-science partnership survey to provide an absolute abundance estimate for anglerfish (see Section 5.2). Five surveys have been carried out to date and these cover the main distribution of the anglerfish fishery. The survey is also considered to have greater spatial coverage for megrim and as such is recommended by WKAGME (2008) as the main source of data of megrim relative abundance for the Northern Shelf. Currently, five years of data are available (20052009) as data from the 2010 survey are not yet available, but the time-series will be updated as soon as this becomes available (summer 2010).

For the five years of survey data available, the sample locations and the density of megrim are illustrated in Figure 5.3.5 as numbers (number per square kilometre) and in Figure 5.3.6, as weight (kilograms per square kilometre). The highest densities of megrim occurred close to the 200 m contour in the northern and western areas, and on the eastern slopes of the Rockall plateau; high densities were also present in the northern North Sea.

The results of the survey are presented in Table 5.3.3. The abundance and biomass time-series are given in Figures 5.3.7 and 5.3.8 respectively. The increase in abundance and biomass on the Northern Shelf from 2005 to 2009 was $39 \%$ and $25 \%$ respectively: In each case, over $50 \%$ of this abundance and biomass was contained in Subarea IV (North Sea).

Using the ration of the average abundance estimate from the first two years of the time-series with the last two years in line with the method proposed by the EC for setting TACs for category six stocks, gives an increase in relative biomass of $25 \%$ for the entire survey area (Table 5.3.4). Split by area, the biomass estimates increase by 28 and $23 \%$ for ICES Area VI and IV (partial coverage) respectively. It should be noted that the confidence intervals, particularly for IV are very broad (Figure 5.3.8) and it is not possible to say with any certainty whether this increase is significant. The confidence estimates for Area VI are narrower and the trend shown in Figure 5.3.7 would indicate that biomass has increased, but has levelled out in the latter part of the timeseries.

## Commercial cpue

Logarithmic lpues for Scottish, French and Irish vessels split by mesh bands corresponding to gear groups TR1 ( $>100 \mathrm{~mm}$ ) and TR2 ( $>70<100 \mathrm{~mm}$ ) as defined by 1342/2008 are available for both VIa and IV (France and Scotland only) (the last available year) based on data presented to SGMOS 09-05 (Part 2) and from 2003 to 2009 for Ireland (VIa only). These are presented in Figure 5.3.9.The commercial lpues are also contrasted with the anglerfish logarithmic cpues for comparison. Between 2005 and 2007, both the commercial lpues and the survey cpues trends are reasonable consistent across fleets with all showing generally positive increases, with the exception of the Irish TR2 fleet. It should be noted that the IRE TR2 fleet has been discontinued due to the prohibition of mesh sizes $<120 \mathrm{~mm}$ for vessels targeting fish (EC regulation 43/2008). Since 2007, the lpues for both the SCO TR1 and FR TR1 fleets show a dramatic increase as has the IRE TR2 since 2008. These signals give a much stronger positive signal than the survey-series during this period. It is not possible to determine how much this could be attributed to changes in megrim abundances or changes in targeting behaviour, but there reasons to suspect that there has been significant changes in targeting behaviour. Over the period, there have been reduced fishing opportunities for other species (e.g. cod) and reduced effort allocations inside the West of Scotland management line, particularly affecting Scottish and Irish vessels; this may have resulted in increased targeting of anglerfish and megrim to the west of the management line, where effort opportunities are far less constrained.

Logarithmic lpues for two Scottish commercial fleets (SCO TR1 and SCO TR2) in Area IV from 2003 to 2008 are given in Figure 5.3.10. These are also contrasted with the log transformed indices from the anglerfish survey. The trends between the two commercial lpue indices are consistent and show a positive trend during the past few years. However, the survey cpue is more variable and doesn't appear to follow the commercial lpues. Care should be taken in interpreting the commercial lpue's given possible shifts in targeting behaviour.

Lpue data for Division VIb is only available for Irish vessels. In 2009, lpue has declined steadily between 2004 and 2008, but has increased marginally in 2009 (Figure 5.3.11) .Irish effort has also continued to decline in recent years and current OTB effort in kW days has declined by $58 \%$ since 2003. No effort data is available for either France of UK, who combined have $76 \%$ of the TAC.

### 5.3.3 Historical stock development

No analytical assessment has been agreed for this stock since 1999.

## State of the stock

The state of the stock is unknown.

### 5.3.4 Short-term projections

There is no accepted analytical assessment for this stock.

### 5.3.5 Biological reference points

## Precautionary approach reference points

No precautionary reference points have been defined for this stock.

## Yield-per-recruit analysis

It was not possible to define $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ values for this stock due to the lack of international catch-at-age data and recent changes in fleet selectivity due to likely changes in targeting behaviour and recent changes in mesh selectivity, which, if fully implemented, will result in a significant change in age selectivity of the gear.

### 5.3.6 Uncertainties and bias in assessment and forecast

There is no accepted analytical assessment for this stock.

### 5.3.7 Recommendation for next Benchmark

In its porposals for benchmarks for 2010 and 2011, ICES notes that for Megrim in Subareas VI and IV:

Could be considered at the same time as megrim in Divisions VII-k and VIIIabd. However, there may still remain major issues with data. Many of the data and assessment problems are similar for Megrim in VI and IV i.e. discards, tuning fleets, consistency in the catch-at-age data, poorly known migrations, complicated life histories. It would be very important to benchmark the assessments in both areas at the same time.

In advance of a benchmark in 2011 it is first necessary to construct international catch numbers/weights-at-length and age for the main fleets engaged in the fishery. For megrim in VI, this requires data from Ireland, UK, France and Spain and for megrim in IV, from the UK. Effort data for the main fleets engaged in both the VI and IV megrim fisheries are required to provide a time-series of trends in commercial lpue in both VI and IV. Progress should be reviewed by WGCSE 2010 at which point the viability of a benchmark process in 2011 should be confirmed.

## Data requirements

International landings numbers and weights-at-length and age for the main fleets engaged in the fishery. For megrim in VI, this requires data from Ireland, UK, France and Spain and for megrim in IV, from the UK.

Effort data for the main fleets engaged in both the VI and IV megrim fisheries are required to construct potential commercial tuning fleets in both VI and IV.

Fishery independent survey indices disaggregated by sex for quarter 1 and 4 surveys in Subdivisions VI and IV.

### 5.3.8 Management considerations

The TAC in VI has not been fully utilised. However, the uptake rate is country specific, with full uptake being reported by some member states. Partial quota by indi-
vidual member states may be an artefact of reduction in effort rather than reflective of a reduction in biomass. Data from the anglerfish survey indicates similar abundance in 2009 as observed in 2010 in Subdivision VI. The TAC in IV has been fully utilised and the data from the anglerfish survey indicate a decrease in biomass in the last year of the time-series, although confidence bands are large. Data from the 2009 survey provides a five year times-series, which is now sufficiently long to apply the EC communication to ICES (Comm 2009 224) regarding Category six stocks.

## References

Laurenson, C. and MacDonald, P. 2008. Collection of fisheries and biological data on megrim in ICES Subarea IVa. Scottish Industry Science Partnership Report No 05/08.

Table 5.3. 1. Megrim in Subarea VIa (upper) and VIb (lower). Nominal catch (t) of Megrim West of Scotland and Rockall, as officially reported to ICES and WG best estimates of landings.

| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| Denmark | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| France | 398 | 455 | 504 | 517 | 408 | 618 | 462 | 192 | 172 | 0 | 135 | 252 | 79 | 92 | 50 | 48 | 53 | 104 | 92 |  |
| Ireland | 317 | 260 | 317 | 329 | 304 | 535 | 460 | 438 | 433 | 438 | 417 | 509 | 280 | 344 | 278 | 156 | 221 | 191 | 172 | 188 |
| Netherlands | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| Spain | 91 | 48 | 25 | 7 | 1 | 24 | 22 | 87 | 111 | 83 | 98 | 92 | 89 | 98 | 45 | 69 | 52 | 5 | 149 |  |
| UK - Eng+Wales+N.Irl. | 25 | 167 | 392 | 298 | 327 | 322 | 156 | 123 | 65 | 42 | 20 | 7 | 14 | 13 | 17 | 10 | 0 | 8 | 6 |  |
| UK - Scotland | 1093 | 1223 | 887 | 896 | 866 | 952 | 944 | 954 | 841 | 831 | 754 | 770 | 643 | 558 | 469 | 269 | 336 | 658 | 868 | 953 |
| UK |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Offical Total | 1924 | 2154 | 2125 | 2047 | 1907 | 2451 | 2044 | 1795 | 1622 | 1394 | 1424 | 1630 | 1105 | 1105 | 859 | 552 | 662 | 966 | 1287 | 1131 |
| Unallocated | 286 | 278 | 424 | 674 | 786 | 1047 | 2010 | 1477 | 1083 | 1254 | 823 | 843 | 723 | 537 | 469 | 9 | 213 | n/a | 8 |  |
| As used by WG | 2210 | 2432 | 2549 | 2721 | 2693 | 3498 | 4054 | 3272 | 2705 | 2648 | 2247 | 2473 | 1828 | 1642 | 1328 | 561 | 875 | 1301 | 1545 |  |
| Area Misreported landings | 339 | 338 | 466 | 735 | 871 | 1126 | 2062 | 1556 | 1156 | 1066 | 868 | 829 | 731 | 544 | 421 | n/a | 212 | 478 | 250 |  |
| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| Belgium | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |


| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | - | - | - | - | - | - | - | - | - | . | 4 | $<0.5$ | $<0.5$ | - | - | - | - | - |  |  |
| Ireland | 196 | 240 | 139 | 128 | 176 | 117 | 124 | 141 | 218 | 127 | 167 | 176 | 87 | 83 | 43 | 68 | 95 | 87 | 68 | 48 |
| Spain | 363 | 587 | 683 | 594 | 574 | 520 | 515 | 628 | 549 | 404 | 427 | 370 | 120 | 93 | 71 | 88 | 59 | 19 | 84 |  |
| UK - Eng+Wales+N.Irl. | 19 | 14 | 53 | 56 | 38 | 27 | 92 | 76 | 116 | 57 | 57 | 42 | 41 | 74 | 42 | 19 | 9 | . |  |  |
| UK - England \& Wales | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . |  | . | . | 1 |  |
| UK - Scotland | 226 | 204 | 198 | 147 | 258 | 152 | 112 | 164 | 208 | 278 | 309 | 236 | 207 | 382 | 372 | 207 | 181 | . | 141 | 178 |
| UK |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Offical Total | 804 | 1045 | 1073 | 925 | 1046 | 816 | 843 | 1009 | 1091 | 866 | 964 | 824 | 455 | 632 | 528 | 382 | 344 | 106 | 294 | 226 |

Unallocated

Table 5.3.2. Megrim in Subarea IV and IIa. Nominal catch (t) of Megrim North Sea, as officially reported to ICES and WG best estimates of landings.

| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 4 | 3 | 2 | 7 | 2 | 7 | 5 | 3 | 5 | 4 | 10 | 2 | 5 | 3 | - | - | 2 | 6 | 3 | 2 |
| Denmark | 2 | 1 | 4 | 6 | 1 | 2 | 7 | 5 | 18 | 21 | 29 | 52 | 8 | 11 | 7 | 1 | 6 | 11 | 31 |  |
| France | - | - | 36 | 25 | 27 | 24 | 14 | 16 | 14 | . | 7 | 5 | 6 | 11 | 9 | 3 | 4 | 18 | 21 |  |
| Germany | . | 6 | 3 | 4 | 1 | 2 | 1 | 2 | 4 | 1 | 3 | 1 | - | 2 | 2 | 4 | 7 | 16 | 5 | 4 |
| Germany, Fed. Rep. of | 3 | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . |  |  |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - | - | - | . |  |  |
| Netherlands | 24 | 28 | 27 | 30 | 28 | 26 | 9 | 20 | 30 | 26 | 20 | 11 | 9 | 7 | 11 | 19 | 22 | 20 | 3 | 2 |
| Norway | - | - | - | - | - | - | - | - | - | - | - | - | - | $<0.5$ | $<0.5$ | <0.5 | 1 | 1 | 4 |  |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | . |  |  |
| Sweden | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| UK - Eng+Wales+N.Irl. | 17 | 9 | 47 | 8 | 19 | 44 | 4 | 3 | 5 | 4 | 2 | 2 | 3 | 1 | 1 | 1 | 9 | 17 |  |  |
| UK - England \& Wales | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | . | 6 |  |
| UK - N. Ireland | . | . | - | . | . | . | . | . | . | . | - | . | - | - | - | - | . | - |  |  |
| UK - Scotland | 1126 | 1169 | 1372 | 1736 | 2000 | 2193 | 3221 | 3091 | 2628 | 2121 | 2044 | 1854 | 1675 | 1235 | 1130 | 958 | 1340 | 1436 | 1526 |  |
| UK |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1476 |
| Official total | 1176 | 1216 | 1491 | 1816 | 2078 | 2298 | 3261 | 3140 | 2704 | 2177 | 2115 | 1927 | 1706 | 1271 | 1160 | 986 | 1391 | 1525 | 1599 | 1753 |
| As used by WG | 837 | 878 | 1025 | 1081 | 1207 | 1172 | 1199 | 1584 | 1548 | 1111 | 1247 | 1098 | 975 | 727 | 739 | n/a | 1179 | 1047 | 1349 |  |
| Area Misreported landings | 339 | 338 | 466 | 735 | 871 | 1126 | 2062 | 1556 | 1156 | 1066 | 868 | 829 | 731 | 544 | 421 | n/a | 212 | 478 | 250 |  |

Table 5.3.3. Estimates of megrim abundance and biomass from Scottish-Irish anglerfish surveys

|  | Abundance (millions) |  |  |  | Biomass (tonnes) |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2005 | 2006 | 2007 | 2008 | 2009 | 2005 | 2006 | 2007 | 2008 | 2009 |
| Area IV (partial) | 11.7 | 11 | 14.8 | 18.9 | 13 | 4652 | 3629 | 5509 | 6953 | 4361 |
| Area VI | 5.5 | 9.3 | 13.8 | 14.4 | 15 | 2444 | 3127 | 4258 | 4063 | 4321 |
| Northern Shelf (partial) | 17.2 | 20.3 | 28.6 | 33.3 | 28 | 7096 | 6757 | 9766 | 11016 | 8673 |

Table 5.3.4. Changes in relative megrim abundance and biomass from Scottish-Irish anglerfish surveys based on percentage changes in mean abundance and biomass from the first three years of the survey relative to the mean of the last two years.

|  | Abundance |  | Biomass |  | Abundance | Biomass |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean 05-07 | Mean 08/09 | Mean 05/07 | Mean 08/09 | Percentage Change |  |
| Area IV (partial) | 12.50 | 15.95 | 4597 | 5657 | $28 \%$ | $23 \%$ |
| Area VI | 9.53 | 14.70 | 3276 | 4192 | $54 \%$ | $28 \%$ |
| Northern Shelf (partial) | 22.03 | 30.65 | 7873 | 9845 | $39 \%$ | $25 \%$ |



Figure 5.3.1. Fishing effort in ICES Division VIa for Irish, French and Scottish vessels by mesh category.


Figure 5.3.2. Change in selection profile of megrim associated with the increase in mesh size from 100 mm (left) to 120 mm (right) associated with the introduction of emergency measures in VIa (EC regulation $43 / 2009$ ).


Figure 5.3.3. Irish Otter trawl effort in ICES Subdivision VIb (Rockall) expressed in both effort hours and kw.days.




Figure 5.3.4. International megrim landing by ICES statistical rectangle for ICES Divisions VIa, VIb and IVa for 2007-2009. Note that no French data was available for 2009.


Figure 5.3.5 Maps of the northern continental shelf around the British Isles showing the number density of megrim caught during the anglerfish surveys 2005-2009. Each circle (blue for Scottish surveys; green for Irish surveys) is centred on the sample location and the size of the circle is proportional to the number density in $\mathbf{n} / \mathrm{km} 2 \mathrm{according}$ to the legend (top left). The red lines indicate the position of the borders between the main ICES sub-areas (labeled with Roman numerals).


Figure 5.3.6. Maps of the northern continental shelf around the British Isles showing the weight density of megrim during the anglerfish surveys 2005-2009. Each circle (blue for Scottish surveys; green for Irish surveys) is centred on the sample location and the size of the circle is proportional to the weight density in $\mathbf{~ k g} / \mathbf{k m}^{2}$ according to the legend (top left). The red lines indicate the position of the borders between the main ICES subareas (labelled with Roman numerals).


Figure 5.3.7. Relative change in ICES Division VI megrim biomass with $95 \%$ confidence intervals from the 2005-2009 anglerfish survey.


Figure 5.3.8. Relative change in ICES Division IV (partial coverage) megrim biomass with 95\% confidence intervals from the 2005-2009 anglerfish survey.


Figure 5.3.9. Change in commercial Log lpue and survey (Area VI) cpue relative to long-term average for Megrim in VIa (Area VI for survey).


Figure 5.3.10. Change in commercial Log lpue and survey cpue relative to long-term average for Megrim in IV.


Figure 5.3.11. Megrim lpue trends for Irish otter trawlers from ICES Subdivision VI.

### 6.1 Irish Sea overview

There is no overview.

### 6.2 Cod in VIIa

Type of assessment
This is an update assessment. The assessment has not yet been included in ICES benchmarking process.

## ICES advice applicable to 2009

"Because the existing recovery plan does not include the elements or measures necessary to rebuild the stock at the current SSB (well below Blim), ICES continues to advise on exploitation boundaries in relation to precautionary limits and recommends that the fisheries for cod be closed until an initial recovery of the cod SSB has been proven. Any catches that are taken in 2008 will prolong the recovery to Bpa."

## ICES advice applicable to 2010

"ICES has evaluated the long-term management plan and found it not precautionary. ... ICES continues to advise on exploitation boundaries in relation to precautionary limits and recommends that the fisheries for cod be closed until an initial recovery of the cod SSB has been proven. Any catches that are taken in 2010 will prolong the recovery to Bpa."

### 6.2.1 General

## Stock description and management units

The stock and the management unit are both ICES Division VIIa (Irish Sea).

## Management applicable to 2009 and 2010

TACs and quotas set for 2009

| Species: Cod <br> Gadus moriua |  | Zone: | VIla <br> (COD/07A.) |
| :---: | :---: | :---: | :---: |
| Belgium | 12 |  |  |
| France | 33 |  |  |
| Ireland | 592 |  |  |
| The Netherlands | 3 |  |  |
| United Kingdom | 259 |  |  |
| EC | 899 |  |  |
| TAC | 899 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TACs and quotas set for 2010

| Species:Cod <br> Gadus morhua | Zone:VIla <br> (COD/07A.) |  |
| :--- | ---: | :--- | :--- |
| Belgium | 9 |  |
| France | 25 |  |
| Ireland | 444 |  |
| The Netherlands | 2 |  |
| United Kingdom | 194 |  |
| EU | 674 | Analytical TAC |
| TAC | 674 |  |

Management of cod is by TAC, days-at-sea limits and technical measures. Technical regulations in force in the Irish Sea, including those associated with the cod recovery plan since 2000, are described in Section 6.1.

## Fishery in 2009

Landings of cod in 2009 (Table 6.2.1) were the lowest recorded. The percentage landed into Northern Ireland remained at approximately $80 \%$ (Table 6.2.2), the majority taken by whitefish otter trawlers and Nephrops trawlers. The percentages landed into southern Ireland, Belgium and UK (England and Wales) in 2009 also remained very similar to 2008 values at roughly $13 \%, 4 \%$ and $3 \%$ respectively. Irish fleets experienced elevated catch rates of cod in the Celtic Sea off SE Ireland in 2009, and 193 tonnes of cod landings incorrectly reported as taken in VIIa were re-allocated to the Celtic Sea. WG landings figures in 2009 were $52 \%$ of the TAC, and have been at around $50 \%$ of the TAC since 2004.

### 6.2.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1 in the WGCSE Report.

## Fishery landings

The input data on fishery landings and age compositions are split into three periods (Figure 6.2.4):

1) 1968-1990. Landings in this period, provided to ICES by stock coordinators from all countries, were assumed to be accurate and were used directly as the input data for the assessment.
2 ) 1991-1999. TAC reductions in this period caused substantial misreporting of cod landings into several major ports in one country, mainly species misreporting. Landings into these ports were estimated based on observations of cod landings by different fleet sectors during regular port visits (see Stock Annex). For other national landings, the WG figures provided to ICES stock coordinators were used, as in period (1).

3 ) 2000 onwards. Cod recovery measures were considered to have caused greater problems with estimation of fishery removals than in period (2). The ICES WG landings data provided by stock coordinators for all countries, as in period (1) were input to B-Adapt and the annual total removals (in excess of the assumed M ) were estimated within the assessment model.

The annual numbers-at-age landed, total landed weight, and the mean weights-at-age in the landings by age class, are given in Tables 6.2.2-6.2.4 and Figures 6.2.1-6.2.4. Previous WG's have shown there are no long-term trends in catch weights-at-age from 1982 onwards. However, weights-at-age prior to 1982 are fixed at constant values lower than estimated for subsequent years, leading to sums-of-products errors, and weights-at-ages $7+$ are becoming patchy for the last few years (Figure 6.2.1). Given these problems, and the likelihood of further deterioration in the quality of the older aged fish, revision of historical catch-at-age data and associated weights is needed.

The catch-at-age data were screened using separable VPA (reference age 3; terminal F $=1.5 ; \mathrm{S}=1.0$; default year and age weighting). The data continue to show a persistent change in residuals for $\log$ catch ratios at ages 1-2 after 1991 (Figure 6.2.5). Outliers at age 5-6 in 2003/2004 and age 1-2 in 2006/2007 are not associated with any obvious anomalies in any national dataset and reflect small catches and sample sizes.

## Discards data

No discards data are included in the assessment. Suitable discards estimates are not available prior to the mid-1990s and are not complete for many subsequent years. Available data indicates that discarding has historically been mainly a function of MLS and therefore mainly restricted to catches of $<1-\mathrm{gp}$ cod. This pattern continued in 2009 for the Irish and UK (NI) data, but although 11 trips were sampled on UK (E\&W) vessels no cod were caught to sample (Figures 6.2.6 and 6.2.7 and Table 6.2.5). Historical F and recruitment for 1-gp cod are therefore underestimated, but it has not been possible yet to compile a matrix of international fleet-raised discards estimates by year and age for use in assessments. Discards data should be fully evaluated in any future benchmark assessment.

## Biological data

The assessment uses constant values of $\mathrm{M}=0.2$ (all ages) and combined-sex proportion mature values of 0 at age 1, 0.38 at age 2 and 1.0 for older ages (see Stock Annex for derivation).

## Survey data used in assessment

The surveys used in the assessment are described in the Stock Annex, and the series are updated in Table 6.2.6.

## Internal consistency of survey data

The survey data during spring each year are of critical importance for tuning the BAdapt and estimating catch bias because adult cod are better represented than during the autumn surveys. The data for these surveys were screened by fitting the SURBA model using settings described in the Stock Annex, and examining the diagnostic plots. The NIGFS-Mar and ScoGFS-1Q surveys do not exhibit any marked yeareffects, and appear to track year-class variations with good consistency (Figure 6.2.8). Strong positive residuals at age 1 are noted for 1994-1996 in the SURBA model fit for NIGFS-Mar (Figure 6.2.8, bottom panels).

## Consistency between survey-series

The three series of summer-autumn 0-gp indices used in the update B-Adapt assessment do not consistently follow the trends in year-class effects from the SURBA model applied to the NIGFS-Mar and ScoGFS-1Q data (Figure 6.2.9). Whilst the surveys give similar signals for some year classes, there are some years (e.g. 2004-2005 and 2001) where the series diverge noticeably. The NIGFS-Mar and ScoGFS-1Q SURBA models provide very similar trends in year-class strength.

## Commercial cpue

Commercial cpue data are available for this stock but are not currently used in the assessment.

## Other relevant data

Table 6.2.6 includes indices of abundance from the UK Fisheries Science Partnership (www. cefas.co.uk/fsp). These are not used in the update assessment and have not yet been evaluated through any benchmarking process, although are presented as supporting evidence (WD 10). The SSB trends from the UK Fisheries Science Partnership trawl surveys support the trends given by the NIGFS-1Q survey from 2004 onwards (Figure 6.2.10).

A Cefas Q4 IBTS trawl survey-series covering the Irish Sea and Celtic Sea in November commenced in 2004. Cod abundance indices will be provided from this survey in future.

The latest in a series of cod SSB estimates from applications of the annual egg production method, using gene probes to identify early-stage cod eggs, are available for 2008 (WD 11). Further estimates will become available from surveys in 2010. These will be evaluated in future benchmark assessments.

### 6.2.3 Historical stock development

## Deviations from Stock Annex

The assessment does not deviate from the procedure used last year and described in the Stock Annex.

## Software used and model options chosen

The B-Adapt method is described in the Stock Annex. Software version B-AdaptF.exe (13/5/06) was used to allow estimation of removals bias from 2000 onwards.

Model settings for the update assessment are given in Table 6.2.7. B-Adapt can use survey data for the year after the last year of catch data, and in this assessment the
survey indices for NIGFS-Mar in 2010 are used. An input F-multiplier for 2010 is required for adjusting the survey indices to the start of the year. In view of the new cod recovery measures which involved a $25 \%$ reduction in cod TAC in 2010, an Fmultiplier of 0.75 was applied in 2010.

## Input data types and characteristics

New data added to the update B-Adapt assessment are the fishery landings data for 2009, the NIGFS-Mar survey data for 2010 and the NIGFS-Oct, UK (BTS-3Q) and NIMIK 0-gp indices for 2009. The update B-Adapt assessment follows the same procedure as in the 2009 assessment by including the sample-based estimates of landings at three major ports from 1991-1999, whilst estimating removals in excess of the assumed natural mortality rate in subsequent years. The sample based estimates of landings for 2000-2002 and 2005 provide a comparison with the B-Adapt removals estimates.

## Data screening

Screening of input catch and survey data is described in Section 6.2.2.

## Final update assessment: diagnostics

The diagnostics of the update B-Adapt run are given in Table 6.2.8. Note that these are from the non-bootstrap application of the model. The catchability residuals from the update assessment are given in Figure 6.2.11. A trend in catchability residuals for $2-4$ year old cod exists in the first five years of the NIGFS-Mar survey-series. This is not reflected in the SURBA residuals shown in Figure 6.2.8. In contrast, the three positive values at age 1 in 1994-1996 in NIGFS-Mar B-Adapt residuals are evident in the SURBA analysis, indicating a change in survey selectivity.

## Final update assessment: Retrospective analysis

The estimation of catch bias in B-Adapt effectively removes survey catchability trends from 2000 onwards, and the assessment therefore exhibits no retrospective bias (Figure 6.2.13).

## Final update assessment: long term trends

The population numbers and F at age from the update B-Adapt assessment are given in Tables 6.2.9 and 6.2.10, and the VPA summary data are given in Table 6.2.11. These are the point estimates from the non-bootstrap option. The long-term trends in landings, F, SSB and recruitment are shown in Figure 6.2.14, using the bootstrap option to give 5 th and 95 th percentiles from 1000 boot-strap runs selecting randomly from the survey catchability residuals. Note that the 50th percentiles differ slightly from the point estimates from the non-bootstrap option.

The B-Adapt estimates of total removals for 2000-2009 (in excess of the WG landings figures and natural mortality $\mathrm{M}=0.2$ ) may represent unaccounted discards, landings and additional natural mortality. The B-Adapt estimates of total removals (including unaccounted removals) were close to the WG landings figures including samplebased estimates for 2000 and 2001, but the $90 \%$ confidence limits of the B-Adapt estimates for 2002 and 2005 lie just above the WG landings estimates.

The recruitment trends from B-Adapt are very similar to the indices from SURBA for the NIGFS-Mar and ScoGFS-1Q surveys (Figure 6.2.15), indicating that the historical trends are well captured by the survey and fishery age-composition data. The SURBA and B-Adapt indices of SSB indicate very low SSB since 2005, and continued high to-
tal mortality rates. Given the highly truncated age composition in the stock, and the internal procedure in SURBA for estimating recent $Z$, the SURBA trends in $Z$ are probably poorly estimated.

In order to investigate the sensitivity of this assessment to the B-Adapt estimates of total removals, another assessment was conducted using the same software and settings, but without estimating the bias. Figure 6.2 .12 presents the results. Although the values of SSB and recruitment are lower without the estimated additional removals, both assessment runs indicate that recent SSB and recruitment both have been at historic lows in recent years. Trends in Fbar are reasonably consistent between the model runs.

## Comparison with previous assessments

The retrospective analysis (Figure 6.2.13) provides a comparison with the results of the assessment carried out in 2009. The current assessment is a direct update without any changes to procedures or data. The current assessment is very consistent with the previous assessment.

## The state of the stock

The spawning-stock biomass has declined ten-fold since the late 1980s and is suffering reduced reproductive capacity ( $\mathrm{SSB}<\mathrm{Blim}$ of 6000 t ).

The fishing mortality estimates since 1988 have remained above the $\mathrm{F}_{\text {lim }}$ value of $\mathrm{F}=1.0$ and the stock has therefore been harvested unsustainably over this period.

Fishing mortality throughout the assessment period has been well above the candidate reference points ( $\mathbf{F}_{\max }$ and $\mathbf{F}_{0.1}$ ) associated with high long-term yields and a low risk of depleting the productive potential of the stock.

Recruitment has been below average for the past seventeen years. The 2002 to 2008 year classes are amongst the smallest on record and all lie below a segmented regression line fitted to the stock-recruit data, indicating lower than expected recruitment given the SSB estimates (Figure 6.2.16). The 2009 data show increased recruitment compared the recent period of poor recruitment, but still below the long-term average. Preliminary indications suggest the 2009 year class to be of similar magnitude to the 2000-2001 year classes. This recruitment is some way above the segmented regression line on Figure 6.2.16, and thus indicates a higher than expected recruitment given the estimated SSB. The estimated breakpoint in the regression is close to the $B_{p a}$ of 10000 t .

### 6.2.4 Short-term predictions

Due to the inability to identify the source of the bias in removals estimates from BAdapt assessment, and the relationship between future TACs and total removals, detailed short-term catch forecasts have not been given for this stock for several years. The update B-Adapt assessment, including a $25 \% \mathrm{~F}$ reduction in 2010, indicates a $300 \%$ increase in SSB between 2010 and 2011. This is a consequence of the high 2009 recruitment entering the spawning stock, after numerous years of very weak recruitment.

| SSB percentile | 2008 |  | 2009 | 2010 |
| :--- | :---: | :---: | :---: | :---: |
| 2011 |  |  |  |  |
| 5th | 1262 | 889 | 500 | 1258 |
| 25th | 1518 | 1083 | 631 | 1821 |
| 50th | 1728 | 1238 | 744 | 2249 |
| 75th | 1981 | 1418 | 873 | 2874 |
| 95th | 2363 | 1685 | 1083 | 4028 |

### 6.2.5 Medium-term projections and MSY evaluation

## Medium-term projections

Medium-term projections are carried out to look at the possible future trends in the stock in response to changes in total mortality. The contribution of the fishery to the total removals estimates over and above reported landings is unknown.

## Estimating recruiting year-class strength

Following the recommendation from RGNSDS (2007) that bootstrapping the 19922006 recruitment estimates may have led to overoptimistic forecasts, 2002 was chosen as the starting year for this assessment's medium-term projections.

The stock-recruit plot (Figure 6.2.16) shows that from 2002 to 2008 the recruitment estimates were well below the segmented regression line, but well above it in 2009. As we do not yet know whether the increase in 2009 is the start of a period of higher recruitment or a one-off high value, two sets of forecasts have been run, one using the year range 2002 to 2009 (including the higher 2009 value) and the other being a more pessimistic view using only the lower (and more recently typical) values seen between 2002 to 2008.

|  |  |  | Number at age 0 <br> (‘000) |
| :---: | :---: | :--- | :---: | :---: |
| Year | Year class | Source | 384 |
| 2007 | 2007 | B-Adapt (point estimate) | 574 |
| 2008 | 2008 | B-Adapt (point estimate) | 3742 |
| 2009 | 2009 | B-Adapt (point estimate) | $1192^{1}$ |
| 2010 | 2010 | Bootstrap 2002-2008 y.c.: (50th percentile) | 1040 |
|  |  | GM (2002-2008 y.c.) | $1262^{1}$ |
| 2010 | 2010 | Bootstrap 2002-2009 y.c.: (50th percentile) | 1221 |
|  |  | GM (2002-2009 y.c.) |  |

${ }^{1}$ Average of $50^{\text {th }}$ percentiles over 10-year B-Adapt projection

## Scenarios examined

The mortality rate due to removals in excess of the assumed natural mortality of $\mathrm{M}=0.2$ is referred to below as $\mathrm{F}^{*}$. Four medium-term stochastic projections were carried out using the bootstrap option in B-Adapt:

1 ) Zero $\mathrm{F}^{*}$ from 2010 onwards with recruitment estimated from model estimates for the year classes observed from 2002-2008.
2 ) Zero $\mathrm{F}^{*}$ from 2010 onwards with recruitment estimated from model estimates for the year classes observed from 2002-2009.
3 ) $25 \%$ reduction in $F^{*}$ per year until $F^{*}$ attains the value of $\mathrm{F}=0.4$ adopted by the Commission as the long-term management objective. Recruitment es-
timated from model estimates for the year classes observed from 20022008.

4 ) $25 \%$ reduction in $\mathrm{F}^{*}$ per year until $\mathrm{F}^{*}$ attains the value of $\mathrm{F}=0.4$ adopted by the Commission as the long-term management objective. Recruitment estimated from model estimates for the year classes observed from 20022009.

Projections 3 and 4 represent annual reductions in $F^{*}$ equivalent to reductions in $F$ that Management Plan may seek to achieve through annual $25 \%$ reductions in TAC. However, the bootstrap procedure does not simulate any additional variability and risk associated with limits on inter-annual TAC variability, or any changes in discarding or compliance.

The removals figures generated in the projection implicitly include the level of removals bias estimated by B-Adapt for 2009 in each simulation. It is currently not possible to attribute these to any actual losses not accounted for in the model inputs, or to any remaining bias due to incorrect assumptions in the B-Adapt implementation.

## Model inputs

Model inputs were as follows:

- Number of simulations: 1000:
- Recruitment from 2010 onwards: bootstrapped in each simulation from model estimates for the year classes described in the scenarios examined section.
- Status quo F: B-Adapt F(2-4) for 2009 in each simulation.
- Intermediate year assumption: To allow for a potential reduction in $\mathrm{F}^{*}$ in 2010 associated with the $25 \%$ TAC reduction, an F-multiplier of 0.75 was applied in 2010.


## Results

Reducing $\mathrm{F}^{*}$ to zero from 2010 onwards allows a high probability of recovery of SSB to above $\mathrm{B}_{\mathrm{pa}}$ by 2015 (Figures 6.2.17 and 6.2.18) regardless of recruitment scenarios, with the more optimistic $02-09$ recruitment model having $99 \%$ probability compared to $98 \%$ probability for the $02-08$ recruitment.

A stepwise reduction in $\mathrm{F}^{*}$ by $25 \%$ per year (until the year when the 50th percentile of $\mathrm{F}^{*}$ reaches 0.40 ) is more dependent on the recruitment range used in the model (Figures 6.2.19 and 6.2.20). The $02-08$ range results in a $26 \%$ probability of $\mathrm{SSB}>\mathrm{B}_{\lim }$ by 2015 and zero probability of achieving $B_{p a}$, with the $02-09$ range showing a $46 \%$ probability of SSB > Blim by 2015 and $10 \%$ probability of achieving $B_{p a}$.

## MSY evaluations

The results for the MSY evaluation are presented in Table 6.2.12 and Figures 6.2.216.2.25. Given the uncertainty in the $F$ estimation for the most recent years in this stock, only data up to 2005 was used. The AIC values are similar for each curve, the smooth hockey stick has the lowest value but the difference between values is small relative to the scale and there is no clear appropriate model selection based on this statistic.

The Fcrash values for Ricker (1.63) and Beverton-Holt (1.68) are very similar with the Hockey stick model showing a lower value (1.21), and all values are some way above
the current $\mathrm{F}_{\lim }$ (1.0). The lower 5th percentile of the Fcrash distribution is very close to $\mathrm{F}_{\mathrm{pa}}(0.72)$ for the Hockey stick ( 0.74 ), but higher for the Ricker ( 0.91 ) and BevertonHolt (0.88).

For both the Ricker and Beverton-Holt curves, one of the ADMB parameters is welldefined ( $13 \%$ c.v.), while the other is poorly determined ( $41-43 \%$ c.v.). In the case of the Beverton-Holt curve, both unscaled parameters are very poorly determined ( $>500 \%$ c.v.s), highlighting the importance of re-parameterisation of the usual Bever-ton-Holt formulation prior to estimation. In contrast, both smooth hockey stick parameters are well determined with c.v.s of $14 \%$ and $20 \%$. Although $\mathrm{B}_{\text {msy }}$ levels are very poorly determined in all three cases (c.v.s ranging from $130 \%$ to $480 \%$ ), and MSY in the case of Ricker and Beverton-Holt (c.v.s of 110-520\%), the corresponding Fmsy level is reasonably well determined for all three models, with c.v.s ranging from $27 \%$ to $36 \%$, and median values ranging from 0.25 to 0.54 . The Ricker curve provides the highest and most precisely determined $\mathrm{F}_{\text {msy }}$ value, and is also regarded as the most biologically plausible model for cod, given that cod are cannibalistic. However, the fit of the stock-recruit curves (AIC values) and corresponding estimates of precision for parameters (stock-recruit and $\mathrm{F}_{\mathrm{msy}}$ ) does not exclude any of the models from being considered, and $\mathrm{F}_{\text {msy }}$ values from all three models are therefore presented.

## Conclusion

The models used do not included uncertainty due to ecosystem effects and multispecies interactions affecting growth, maturity and natural mortality and therefore the stock trajectory estimated at low fishing mortality rates is considered to be highly uncertain. The assessment for this stock does not currently include discards information, and although the model estimates unallocated removals in excess of natural mortality from the year 2000 onwards, this cannot be attributed to any particular mortality source. The assessment therefore incorporates discard mortality from 2000 onwards, but not before. The available data suggests that discarding of Irish Sea cod is predominantly an effect of the MLS, and thus affects the recruiting fish. Recruit estimates for this stock are therefore uncertain.
$\mathrm{F}_{\text {msy }}$ estimates are reasonably well determined for all three models and these models cannot be distinguished based on the current data. Consequently the definition of $\mathrm{F}_{\text {msy }}$ for the Irish Sea cod stock is dependent on whether it is considered that recruitment will be reduced or either remain constant or continue to increase at high stock abundance; the choice between the Ricker on the one hand and the smooth hockey stick and Beverton-Holt models on the other. The Ricker curve is the most plausible based on biological considerations, but until more data are collated at high stock abundance the recruitment dynamics at high stock abundance will be uncertain.

Consequently a definitive $\mathrm{F}_{\text {msy }}$ value cannot be determined for Irish Sea cod based on the current information. On the basis of the three models that have equally plausible fits to the stock and recruit estimates, a range of $0.25-0.54$ would be considered consistent with $\mathrm{F}_{\mathrm{msy}}$ for Irish Sea cod.

### 6.2.6 Biological reference points

The current precautionary reference points for Irish Sea cod are given below:

| $\operatorname{Blim}$ | 6000 t | $\mathrm{B}_{\mathrm{pa}}$ | 10000 t |
| :--- | :---: | :---: | :---: |
| Flim | 1.00 | $\mathrm{~F}_{\mathrm{pa}}$ | 0.72 |

### 6.2.7 Management plans

The Irish Sea cod management plan, as described in Council Regulation (EC) 1342/2008 was evaluated independently by ICES in 2009 using the approach adopted in AGCREMP 2008 and found to be not consistent with the ICES Precautionary Approach (WGCSE 2009).

The long-term target for the management plan is a fishing mortality of 0.4 , based on the EU-Norway negotiated target for North Sea cod. This target is within Fmsy range for Irish Sea cod, and well below the current estimates of total removals mortality in excess of $\mathrm{M}=0.2$.

### 6.2.8 Uncertainties and bias in assessment and forecast

## Landings data

The quality of the commercial landings and catch-at-age data for this stock deteriorated in the 1990s following reductions in the TAC without associated control of fishing effort. The Working Group has, since the 1990s, attempted to overcome this problem by incorporating sample-based estimates of landings from three major ports in the WG landings figures. The data for this method have become more limited since 2003, and the WG uses the B-Adapt modelling approach to estimate subsequent removals from 2000 onwards. The unaccounted removals figures given by B-Adapt could potentially include components due to increased natural mortality and discarding as well as misreported landings or catches from the stock taken outside VIIa, albeit distributed according to the age composition in the landings.

The French landings data used in 2009 are provisional and subject to change in at the next assessment (Table 6.2.2). This data accounts for less than $1 \%$ of the 2009 landings ( $\sim 3$ tonnes).

## Discarding

Estimates of discards are patchy for Irish Sea cod, although more comprehensive sampling is now required through the EU Data Collection Framework. Discarding has historically been mainly at age 1, and the absence of raised estimates of discarding for all fleets will result in under-estimation of historical F at age 1. Strict controls on catch reporting following the introduction of the Registration of Fish Buyers and Sellers regulations has resulted in documented increases in discarding of cod above the MLS off the west of Scotland and in the Celtic Sea (see Sections 3.2 and 7.2). This could also occur in the Irish Sea, although observer data in 2008-2009 provided no evidence for this. Compliance with catch composition rules for some fleets could also result in increased discarding of cod. Implementation of unbiased sampling schemes to estimate discarding with adequate precision is likely to be of increasing importance for this stock to prevent further deterioration in fishery catch data.

## Surveys

The Irish Sea has relatively good survey coverage up to 2010. The surveys in general give consistent signals of fish abundance-at-age. All survey data except the UK(BTS3 Q ) indicate a severe depletion of the SSB during a 7 -year run of very poor recruit-
ment followed by a larger recruitment in 2009. The UK(BTS-3Q) survey does not show this improved recruitment in 2009, but the data only represent a small area of the Irish Sea and may not be representative of the Irish Sea as a whole. The UK Fish-eries-Science Partnership surveys of the Irish Sea cod spawning grounds in spring 2005-2010 (not in the assessment), carried out using commercial trawlers, indicated a widespread distribution of cod mostly at low density but with some localized aggregations (Figure 6.2.26, WGCSE 2010 WD10). The time-series of SSB indices shows a downward trend similar to that shown by NIGFS-1Q which is used in the assessment (Figure 6.2.10), and the highly truncated age composition of cod in the FSP surveys supports the ICES assessment, indicating continuing high mortality rates. Estimates of cod SSB from applications of the annual egg production method, although slightly higher than the B-Adapt estimates, are still below Blim and show a similar trend in SSB to the assessment (Figure 6.2.27).

## Model formulation

The B-Adapt estimates of removals bias continue to vary around relatively high values of 2.0-3.0 despite more accurate catch reporting and lack of evidence for significant discarding of cod above MLS. There could potentially be unaccounted losses from other sources, for example due to fishery catches taken outside VIIa during seasonal migrations, a gradual shift in distribution to areas beyond VIIa, or increases in natural mortality. The estimates of bias could also be influenced by any remaining non-randomness of survey catchability or outlying values, or by incorrect assumptions in the model (e.g. constant survey catchability, removals bias not agedependent). For this reason, the absolute values of the estimated unallocated removals should not be over-interpreted. There is currently no evidence from surveys and fishery age compositions of a significant improvement in age structure that could be caused by management measures. The interpretation in B-Adapt is that there continues to be a relatively large unaccounted-for removal of fish from the stock, but unfortunately there is currently very little direct evidence to evaluate the potential source(s) of this and how much is due to fishing in VIIa or elsewhere.

## Stock structure and migrations

The VIIa commercial fishery for cod extends into the North Channel, particularly for vessels using mid-water trawls. It is not clear if the cod in this region belong to the Irish Sea stock, the nearby Clyde stock which exhibits dense aggregations of adult fish during spring in the area covered by the Clyde closure, or to other VIa cod populations. Incorrect allocation of catches to stocks could lead to biases in the assessments.

Tagging of cod off Greencastle on the north coast of Ireland (O Cuaig and Officer, 2007), and more limited tagging on UK Fisheries Science Partnership surveys (Armstrong et al., WD2 to WGNSDS 2007), have demonstrated movements of cod between Division VIa and VIIa. Most recaptures in VIIa from cod tagged in VIa have come from the North Channel and in or near the deep basin in the western Irish Sea that is a southward extension of the North Channel. The research surveys used for tuning the VIIa cod assessment cover only the western and eastern Irish Sea, and do not extend into the deeper water of the North Channel, where large catches of cod were made by mid-water trawlers in the 1980s and 1990s.

Recently more Irish Sea cod mark and recapture experiments, and electronic data storage tag (DST) results have been collected and analysed (Bendall et al., 2009). These results show not only spring/summer migrations of cod out of the Irish Sea into the North Channel and VIa, but also migrations south through the deeper channel into
the Celtic Sea. This work is continuing and a further 150 cod have been tagged with DST's in the Irish Sea and Celtic Sea in 2010.

Historical tagging studies have also shown more limited movements of cod between spawning components in the western and eastern Irish Sea, for which the migrations tend to be in a north-south direction. STECF Subgroup SGRST (2005, Appendix 4) concluded that management of the Irish Sea stock on the basis of substock assessment regions would be difficult in practice, particularly the separation of catches when the stock units are mixed. Further tagging and genetics studies are required to investigate stock structure, seasonal movements and mixing in VIIa and neighbouring areas.

### 6.2.9 Recommendations for next benchmark assessment

| Year | Candidate stocks | Supporting justification and comment(s) | Indicated expertise necessary at the benchmark meeting |
| :---: | :---: | :---: | :---: |
| 2012? | Western waters cod stocks (Area VI and VII excl VIId). | Cod stocks in Divisions VI and VII comprise an assemblage of metapopulations with varying degrees of mixing. Fishing effort, predation and other environmental drivers including climate change impact the populations in different ways across the range of the stocks. The stocks have proved difficult to assess due to data deficiencies and an inability to demonstrate responses to changes in fishing effort and other management controls. Improved management advice may benefit more from quantifying the spatial dynamics of cod in relation to spatial variations in fishing and other pressures than by trying to refine the current modelling approaches applied to the current stock definitions and management units. To make progress towards this, an initial Data Workshop is proposed to collate and interpret existing and new data on cod stock structure and mixing, distribution patterns, spatial variations in size/age structure and biological characteristics as well as pressures including predation, fishing and climate. Such analyses will be facilitated by high-resolution spatial data on fishery catches and effort by metier using VMS, rectangle data, employing GIS methods. It will be necessary to develop an international database holding spatially resolved data sets (landings, discards, effort, size/age/biological data, surveys, environmental variables) and data manipulation routines to allow evaluation of the effect on the assessments of altering the stock unit definition. Data on cod movement parameters will be required to allow development of operating models for testing assessment and management procedures and ultimately developing and testing spatially disaggregated assessment models. New data sets e.g. on discarding, biology, predation, surveys and fishing effort/cpue would be evaluated. The Data Workshop would build on and review the outcomes of a major UK collaborative programme on cod stock structure and spatial dynamics, which will be completed in 2011. The ensuing Benchmark Assessment workshop would evaluate the appropriateness of current assessment methods in the light of the Data Workshop outcomes, and explore alternative approaches as candidates for providing management advice. This could potentially include changes to the spatial units for assessment or the development of spatially disaggregated assessment models including mixing coefficients. |  |

### 6.2.10 Management considerations

A number of emergency and cod recovery plan measures have been introduced since 2000 to conserve Irish Sea cod. These include a spawning closure since 2000 and effort control since 2003. There have also been several vessel decommissioning schemes. As it has not been possible to provide analytical catch forecasts in recent years, the TAC has been reduced by $15-20 \%$ annually since 2006 and by $25 \%$ since 2009. These measures may have prevented a further increase in fishing mortality of cod or may have resulted in some reduction in fishing mortality. However, the current assessment does not provide sufficiently robust estimates of fishing mortality to allow the possible changes to be determined.

Although recent recruitment patterns appear well estimated in the assessment, the problem of inaccurate landings and discards estimates makes it difficult to estimate the absolute value and recent trends in fishing mortality. However, all sources of information on age composition in the stock, from the fishery as well as surveys using research vessels and chartered commercial vessels, indicates a continued paucity of cod older than four years of age in the Irish Sea indicating a continued very high mortality rate. Possible causes of this include:

- TACs have not restricted catches as intended. Substantial underreporting of landings is known to have occurred since the 1990s, although there is some indication that this is reduced since 2006. However the assessment continues to indicate a large unaccounted removal of fish. The relative contribution of fishing to this has not been identified;
- The effort reductions have not been sufficient, although considerable effort reductions have been observed in some fleets (particularly vessels using $>100 \mathrm{~mm}$ mesh);
- Cod continues to be taken in mixed demersal fisheries (particularly for haddock, sole and Nephrops);
- Time and area closures have not been sufficient to lead to rebuilding of this stock;
- Other non-fishery causes, such as increased natural mortality, have increased over time.

It is difficult to reconcile the large apparent mortality rate and unaccounted removals in recent years with the reduction in fishing effort by whitefish trawlers (shown by STECF Subgroup SGMOS (2009) and Gerritsen (WD4)), the very low abundance of cod, and the evidence for more accurate catch reporting since the introduction of the Registration of Buyers and Sellers.

The scientific evaluation of the revised cod Management Plan (Council Regulation (EC) $1342 / 2008$ ) indicates that it may not be sufficiently precautionary to allow rebuilding of the Irish Sea cod stock to a level where it can regain historical productivity by 2015 (see WGCSE 2009 Report, Section 9.2). The probability of recovery of the cod stock will be increased by measures to eliminate discards of cod which historically have mainly comprised undersized fish.

A closure of the western Irish Sea spawning grounds for cod from mid February to end of April has been in place since 2000, with an extension to the eastern Irish Sea in 2000. The closure was reviewed in 2007 by STECF SGMOS-07-03. On the basis of the information available, SGMOS-07-03 was unable to determine the extent to which the closure has reduced fishing mortality to a lower value than would otherwise have occurred, through protection of adult cod during spawning or influencing changes in
fishing effort in the different fleets. SGMOS advised that a comprehensive evaluation of how fleet activities have been affected by the closure and other regulations and factors is required to evaluate the cod closure.

Surveys of cod eggs in the Irish Sea in 2008 involving the UK and Ireland indicated that half of the spawning took place in areas not included in the spring spawning closure, indicating that the design of the closure may no longer be optimal (Figure 6.2.26 and WD 11). The spawning closure encompassed most of the spawning in the western Irish Sea although spawning commenced earlier in the east. Preliminary estimates of spawning-stock biomass of cod based on the annual egg production and estimates of fecundity and sex ratio are 2230 t (RSE 43\%) in the western Irish Sea, 2658 t (RSE $25 \%$ ) in the eastern Irish Sea and 4860 t (RSE 18\%) for the whole Irish Sea (Figure 6.2.27). The update B-Adapt assessment provides an SSB estimate of 1801 t for the Irish Sea in 2008, roughly $40 \%$ of the egg production estimate. Although the estimates vary both methods give SSB below Blim, and both indicate drops in SSB from early years values (years 1995 and 2000 for the egg survey). Further estimates of cod SSB from this method will become available in 2011 from surveys carried out in 2010.

## References

Bendall, V. Ó Cuaig, M. Schön, PJ. Hetherington, S, Armstrong, M. Graham, N. Righton, D. 2009. Spatio-temporal dynamics of Atlantic cod (Gadus morhua) in the Irish and Celtic Sea: results from a collaborative tagging programme. ICES CM 2009/J:06, 35 pp .

Table 6.2.1. Nominal landings ( $\mathbf{t}$ ) of COD in Division VIIa as officially reported to ICES, and figures used by ICES.

| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 20091 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 187 | 142 | 183 | 316 | 150 | 60 | 283 | 318 | 183 | 104 | 115 | 60 | 67 | 26 | 19 |
| France | 166 | 148 | 268 | 269 | $\mathrm{n} / \mathrm{a}$ | 53 | 74 | 116 | 151 | 29 | 35 | $18^{2}$ | $17^{2}$ | 3 | - |
| Ireland | 1,414 | 2,476 | 1,492 | 1,739 | 966 | 455 | 751 | 1,111 | 594 | 380 | 220 | $275^{2}$ | $608^{2}$ | 643 | 2485 |
| Netherlands | - | 25 | 29 | 20 | 5 | 1 | - | - | - |  |  |  |  |  |  |
| Spain | - | - | - | - | - | - | - | - | 14 | - | - |  |  |  |  |
| UK (England, Wales \& NI) | 2,330 | 2,359 | 2,370 | 2,517 | 1,665 | 799 | 885 | 1,134 | 505 | 646 | 594 | $591{ }^{2}$ | $423{ }^{2}$ | 545 | 389 |
| UK (Isle of Man) | 22 | 27 | 19 | 34 | 9 | 11 | 1 | 7 | 7 | 5 | n/a | n/a | n/a | n/a | $\mathrm{n} / \mathrm{a}$ |
| UK (Scotland) | 414 | 126 | 80 | 67 | 80 | 38 | 32 | 29 | 23 | 15 | 3 | $6^{2}$ | $2^{2}$ | n/a | n/a |
| Total | 4,533 | 5,303 | 4,441 | 4,962 | 2,875 | 1,417 | 2,026 | 2,715 | 1,477 | 1,179 | 967 | 950 | 1,091 | 1217 | 656 |
| Unallocated | 54 | -339 | 1,418 | 356 | 1,909 | -143 | 226 | -20 | -192 | -107 | -57 | -110 | -389 | -556 | -188 |
| Total as used by WG | $4587{ }^{3}$ | $4964{ }^{3}$ | 58593 | $5318{ }^{3}$ | $4784{ }^{3}$ | $1274{ }^{4}$ | $2252{ }^{4}$ | $2695{ }^{4}$ | $1285{ }^{4}$ | $1072{ }^{4}$ | $910^{4}$ | $840^{4}$ | $702{ }^{4}$ | $661{ }^{4}$ | 468 |

${ }^{1}$ Preliminary. ${ }^{2}$ Revised. $\quad \mathbf{n} / \mathbf{a}=$ not available ${ }^{3}$ includes sample-based estimates of landings into three ports ${ }^{4}$ based on official data only. ${ }^{5}$ Estimate due to incorrect submission to ICES.

Table 6.2.2. Cod in VIIa. Working Group figures for annual landings by country since 2000.

| (a) WG landings (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | NI | E\&W | Scotland | Ireland | France | Belgium | Isle of Man | Netherlands | Total | TAC | \%uptake |
| 2000 | 638 | 156 | 39 | 321 | 52 | 56 | 11 | 0 | 1273 | 2100 | 61 |
| 2001 | 697 | 209 | 32 | 645 | 361 | 300 | 8 | 0 | 2251 | 2100 | 107 |
| 2002 | 983 | 171 | 39 | 953 | 251 | 294 | 1 | 2 | 2695 | 3200 | 84 |
| 2003 | 381 | 118 | 32 | 415 | 145 | 187 | 7 | 0 | 1285 | 1950 | 66 |
| 2004 | 539 | 103 | 15 | 271 | 37 | 103 | 5 | 0 | 1072 | 2150 | 50 |
| 2005 | 523 | 72 | 4 | 168 | 31 | 108 | 3 | 0 | 910 | 2150 | 42 |
| 2006 | 552 | 32 | 6 | 172 | 17 | 59 | 3 | 0 | 840 | 1828 | 46 |
| 2007 | 396 | 27 | 2 | 191 | 18 | 66 | 2 | 0 | 702 | 1462 | 48 |
| 2008 | 523 | 22 | 1 | 85 | 3 | 27 | 1 | 0 | 662 | 1199 | 55 |
| 2009* | 375 | 15 | 0 | 55 | 3 | 19 | 1 | 0 | 468 | 899 | 52 |

(b) Percentage of annual total

| Year | NI | E\&W | Scotland | Ireland | France | Belgium | Isle of Man | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 50.1 | 12.3 | 3.0 | 25.2 | 4.1 | 4.4 | 0.9 | 0.0 | 100 |
| 2001 | 31.0 | 9.3 | 1.4 | 28.6 | 16.1 | 13.3 | 0.4 | 0.0 | 100 |
| 2002 | 36.5 | 6.4 | 1.5 | 35.4 | 9.3 | 10.9 | 0.0 | 0.1 | 100 |
| 2003 | 29.7 | 9.2 | 2.5 | 32.3 | 11.3 | 14.6 | 0.6 | 0.0 | 100 |
| 2004 | 50.3 | 9.6 | 1.4 | 25.2 | 3.5 | 9.6 | 0.4 | 0.0 | 100 |
| 2005 | 57.5 | 7.9 | 0.5 | 18.5 | 3.5 | 11.8 | 0.3 | 0.0 | 100 |
| 2006 | 65.7 | 3.8 | 0.7 | 20.4 | 2.0 | 7.1 | 0.3 | 0.0 | 100 |
| 2007 | 56.5 | 3.8 | 0.3 | 27.2 | 2.5 | 9.5 | 0.3 | 0.0 | 100 |
| 2008 | 78.9 | 3.4 | 0.2 | 12.8 | 0.5 | 4.0 | 0.2 | 0.0 | 100 |
| 2009* | 80.1 | 3.1 | 0.0 | 11.7 | 0.6 | 4.1 | 0.3 | 0.0 | 100 |

* French data is provisional for 2009.

Table 6.2.3. Cod in VIIa. Landings numbers-at-age used in the update B-Adapt assessment.


Table 6.2.4. Cod in VIIa. Mean weights-at-age in the landings (used for stock and catch).

| Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7+ |
| 1968 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1969 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1970 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1971 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1972 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1973 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1974 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1975 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1976 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1977 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1978 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1979 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1980 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1981 | 0 | 0.61 | 1.66 | 3.33 | 5.09 | 6.19 | 6.76 | 8.3 |
| 1982 | 0 | 1.01 | 1.524 | 3.488 | 5.573 | 7.592 | 8.697 | 10.18 |
| 1983 | 0 | 0.995 | 1.842 | 3.988 | 5.964 | 7.966 | 9.306 | 10.925 |
| 1984 | 0 | 0.679 | 1.813 | 3.808 | 5.865 | 7.475 | 9.818 | 10.748 |
| 1985 | 0 | 0.783 | 2.023 | 4.244 | 5.825 | 7.5 | 8.81 | 9.504 |
| 1986 | 0 | 0.805 | 1.825 | 3.862 | 5.855 | 7.391 | 8.116 | 9.471 |
| 1987 | 0 | 0.713 | 2.161 | 3.91 | 6.41 | 7.821 | 9.888 | 10.658 |
| 1988 | 0 | 0.607 | 1.563 | 3.756 | 5.668 | 8.017 | 9.749 | 10.208 |
| 1989 | 0 | 0.936 | 1.846 | 3.223 | 5.408 | 6.571 | 8.256 | 11.052 |
| 1990 | 0 | 0.842 | 1.938 | 3.572 | 5.277 | 7.531 | 8.398 | 12.699 |
| 1991 | 0 | 0.856 | 1.637 | 3.542 | 5.419 | 6.39 | 8.507 | 10.397 |
| 1992 | 0 | 0.813 | 1.964 | 3.993 | 5.975 | 6.923 | 8.509 | 11.1 |
| 1993 | 0 | 0.847 | 1.706 | 3.666 | 5.675 | 7.365 | 9.486 | 10.761 |
| 1994 | 0 | 0.798 | 1.923 | 3.608 | 6.08 | 7.68 | 8.272 | 11.258 |
| 1995 | 0 | 0.9 | 1.84 | 4 | 5.791 | 8.452 | 8.712 | 9.56 |
| 1996 | 0 | 0.98 | 1.625 | 3.256 | 5.298 | 7.721 | 8.836 | 12.256 |
| 1997 | 0 | 0.846 | 1.937 | 3.624 | 5.291 | 6.115 | 8.672 | 11.263 |
| 1998 | 0 | 0.925 | 1.647 | 3.729 | 5.371 | 7.033 | 8.833 | 12.155 |
| 1999 | 0 | 0.853 | 1.624 | 3.179 | 5.505 | 7.517 | 10.137 | 12.618 |
| 2000 | 0 | 0.851 | 1.985 | 3.573 | 5.138 | 7.148 | 8.528 | 7.692 |
| 2001 | 0 | 0.99 | 1.823 | 4.149 | 5.606 | 7.332 | 8.471 | 9.667 |
| 2002 | 0 | 0.942 | 1.836 | 3.439 | 5.727 | 7.708 | 9.639 | 10.761 |
| 2003 | 0 | 1.205 | 1.662 | 3.287 | 5.425 | 10.198 | 10.308 | 13.696 |
| 2004 | 0 | 1.112 | 2.202 | 3.634 | 6.505 | 7.638 | 8.937 | 7.572 |
| 2005 | 0 | 0.913 | 1.938 | 3.514 | 5.318 | 7.739 | 7.94 | 12.237 |
| 2006 | 0 | 0.826 | 1.843 | 3.666 | 4.709 | 6.393 | 7.562 | 12.236 |
| 2007 | 0 | 0.832 | 1.852 | 3.781 | 5.347 | 7.991 | 10.038 | 0 |
| 2008 | 0 | 0.894 | 1.586 | 3.543 | 6.001 | 7.573 | 9.723 | 8.123 |
| 2009 | 0 | 1.097 | 2.006 | 3.458 | 5.314 | 7.1 | 6.815 | 0 |

Table 6.2.5. Cod in VIIa. Estimates of numbers discarded in 1996-2009. Data are numbers (' 000 fish) discarded by each fleet, estimated from numbers per sampled trip raised to total fishing effort by each fleet, for the range of quarters indicated. Sampling scheme (a) provides independent self-sampling estimates for the UK(NI) Nephrops fishery also covered by observer data in schemes (b) and (d). An asterisk indicates years/fleets where the data are raised to the trip level rather than to the entire fleet.
a) Self sampling scheme: N.Ireland single trawl Nephrops vessels. Estimates are extrapolated to all N.Ireland vessels catching Nephrops (single and twin trawl)

|  | 1996 Q1-4 | 1997 Q1-4 | 1998 Q1-4 | 1999 Q 1-4 | 2000 Q1-4 | 2001 Q1-4 | 2002 Q1-4 | 2003 Q1 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 43 trips | 39 trips | 48 trips | 39 trips | 44 trips | 43 trips | 35 trips | 8 trips | 0 trips | 0 trips | 0 trips | 0 trips |  |  |
| 0 | 56 | 3 | 0 | 70 | 32 | 4 | 0 | 0 |  |  |  |  |  |  |
| 1 | 82 | 63 | 14 | 83 | 397 | 31 | 22 | 0 |  |  |  |  |  |  |

(b) Observer scheme: N.Ireland vessels catching Nephrops (single trawl only) (*not raised to fleet level - no. of fish)

|  | 1996 | 1997 | 1998 | 1999 Q 3-4 | 2000 Q1-3 | 2001 Q1 | 2002 | 2003 | 2004 | 2005 | 2006 Q 3-4* | 2007 Q1-4 | 2008 Q 1-4 | 2009 Q1-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 0 trips | 0 trips | 0 trips | 4 trips | 6 trips | 1 trip | 0 trips | 0 trips | 0 trips | 0 trips | 9 trips * | 29 trips | 55 trips | 30 trips |
|  |  |  |  | 0 | 0 | 0 |  |  |  |  | 19 | 5.0 | 2.5 | 50.0 |
|  |  |  |  | 0 | 53 | 0 |  |  |  |  | 7 | 15.2 | 2.7 | 8.7 |
|  |  |  |  |  |  |  |  |  |  |  |  | 0.6 | 0.7 | 0.3 |
| (c) Observer scheme: N.Ireland midwater trawl |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1996 | 1997 Q 2-4 | 1998 Q 1-3 | 1999 Q 3-4 | 2000 Q1 | 2001 Q1 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| Age | 0 trips | n/a | n/a | 5 trips | 4 trips | 2 trips | 0 trips | 0 trips | 0 trips | 0 trips | 0 trips | 0 trips | 1 trip | 1 trip |
|  |  | 0 | 0 | 1.6 | 0 | 0 |  |  |  |  |  |  | 0 | 0 |
|  |  | 17 | 4 | 0 | 0.8 | 0 |  |  |  |  |  |  | 0.45 | 0.03 |
|  |  | 0.5 | 2 | 0 | 0 | 0 |  |  |  |  |  |  | 0 | 0.03 |

(d) Observer scheme: N.Ireland twin trawl (*not raised to fleet level - no. of fish)


Table 6.2.5. Continued.

|  | 1996 Q1-4 | 1997 Q1-4 | 1998 Q1-4 | 1999 Q1-4 | 2000 Q1-4 | 2001 Q1-4 | 2002 Q1-4 | 2003 Q1-4 | 2004 Q1-4 | 2005 Q1-4 | 2006 Q1-4 | 2007 Q1-4 | 2008 Q1-4 | 2009 Q1-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 8 trips* | 8 trips* | 7 trips* | 4 trips ${ }^{\text {* }}$ | 10 trips* | 2 trips** | 1 trip * | 9 trips ${ }^{\text {* }}$ | 11 trips* | 8 trips ${ }^{\text {* }}$ | 5 trips* | 15 trips** | 18 trips ${ }^{\text {* }}$ | 12 trips |
| 0 | 5 | 301 | 0 | 8 | 2320 | 58 | 124 | D | 3213 | 8268 | 774 | 0 | 0 | 107 |
| 1 | 374 | 333 | 202 | 16 | 798 |  | 176 | D | 287 | 632 | 150 | 691 | 441 | 8 |
| 2 | 6 | 87 | 0 | 0 | 10 |  | 0 | 0 | 598 | 0 | 0 | 0 | 0 | 0 |


| (f) Observer scheme: UK(E\&W) Demersil otter trawd |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1996 | 1997 | 1998 | 1999 | 2000 Q1-2 | 101 Q1,2 | $02 \mathrm{Q13}$ | 03 Q12 | 004 Q1-4 | 2005 Q1,2 | 2006 | 2007 | 2008 | 2009 |
| A8P | 0 trips | 0 trips | 0 trips | 0 trips | 21 trips | 8 trips | 4 trips | 4 trips | 7 trips | 4 trips |  |  |  |  |
| 0 |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | see | nt 1 |  |  |
| 1 |  |  |  |  | 38.91 | 9.21 | 3.43 | 0.6 | 17.71 | 126 |  |  |  |  |
| 2 |  |  |  |  | 0.05 | 4.46 | 0 | 0.62 | 0.81 | 036 |  |  |  |  |


|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 Q1,2 | 2002 Q3A | 2003 Q2 | 2004 Q1-3 | 2005 Q2 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Agr | 0 trips | 0 trips | 0 trips | 0 trips | 0 trips | 8 trips | 3 trips | 2 trips | 7 trips | 1 trip |  |  |  |  |
| 0 |  |  |  |  |  | 0 | 0 | 0 | 0.03 | 0 | see comment 1 |  |  |  |
| 1 |  |  |  |  |  | 3.09 | 0.03 | 0 | 0.24 | 0 |  |  |  |  |
| 2 |  |  |  |  |  | 0.7 | 0 | 0 | 0 | 0 |  |  |  |  |

(h) Otserver stheme: UK(E\&W) Danish anchor scine

commert 1 UK data for 2006-2009 avazthe to WGNSDSMGCSE as length conpositims orty, for contined gears

Table 6.2.6. Cod in VIIa: survey indices. Approximate relative standard errors for age groups used in the assessment are given for UK(NI) groundfish surveys. Years/ages used in assessments are in bold.

| ScoGFS :Scottish spring groundfish survey of the Irish Sea |  |  |  |  | Numbers per 10 Hours Fishing |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | 1-gp | 2-gp | 3-gp | 4-gp | 5-gp | 6-gp | 7+ |
| 1996 | 3 | 31 | 44 | 7 | 9 | 0 | 0 |
| 1997 | 22 | 29 | 15 | 13 | 2 | 0 | 1 |
| 1998 | 5 | 81 | 27 | 5 | 1 | 0 | 0 |
| 1999 | 7 | 33 | 93 | 15 | 5 | 0 | 0 |
| 2000 | 51 | 6 | 11 | 16 | 0 | 1 | 0 |
| 2001 | 28 | 56 | 1 | 1 | 4 | 0 | 0 |
| 2002 | 13 | 18 | 37 | 1 | 1 | 0 | 0 |
| 2003 | 8 | 69 | 18 | 9 | 0 | 0 | 0 |
| 2004 | 8 | 11 | 49 | 0 | 3 | 0 | 0 |
| 2005 | 1 | 25 | 8 | 9 | 1 | 0 | 0 |
| 2006 | 2 | 5 | 11 | 0 | 2 | 0 | 0 |
| ScoGFS :Scottish autumn groundfish survey of the Irish Sea October |  |  |  |  | Numbers per 10 Hours Fishing |  |  |
|  |  |  |  |  |  |  |  |
|  | 0-gp | 1-gp | 2-gp | 3-gp | 4-gp |  |  |
| 1997 | 3 | 28 | 19 | 1 | 2 |  |  |
| 1998 | 0 | 8 | 42 | 5 | 0 |  |  |
| 1999 | 164 | 2 | 24 | 6 | 2 |  |  |
| 2000 | 24 | 136 | 4 | 0 | 0 |  |  |
| 2001 | 0 | 0 | 7 | 0 | 0 |  |  |
| 2002 | 0 | 18 | 15 | 9 | 0 |  |  |
| 2003 | 2 | 0 | 27 | 0 | 0 |  |  |
| 2004 | 2 | 12 | 5 | 5 | 0 |  |  |
| 2005 | 3 | 8 | 25 | 2 | 0 |  |  |

NI-GFS March groundfish survey $\quad$ Numbers per 3-miles (approx. 1-h tow) $\quad$ RSE = approximate relative standard error

| Survey | 1-gp | 2-gp | 3-gp | 4-gp | 5-gp | 6-gp | 7+ | RSE(1gp) | RSE(2gp) | RSE(3gp) | RSE(4gp) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 23.257 | 5.005 | 1.965 | 0.248 | 0.000 | 0.031 | 0.017 | 0.58 | 0.36 | 0.26 | 0.40 |
| 1993 | 1.381 | 6.488 | 0.446 | 0.104 | 0.014 | 0.028 | 0.000 | 0.67 | 0.22 | 0.25 | 0.39 |
| 1994 | 13.804 | 1.097 | 1.203 | 0.084 | 0.014 | 0.000 | 0.000 | 0.48 | 0.35 | 0.21 | 0.35 |
| 1995 | 7.007 | 3.862 | 0.200 | 0.108 | 0.000 | 0.010 | 0.000 | 0.30 | 0.25 | 0.41 | 0.39 |
| 1996 | 11.061 | 3.293 | 1.117 | 0.014 | 0.088 | 0.000 | 0.013 | 0.62 | 0.18 | 0.21 | 1.00 |
| 1997 | 5.373 | 4.158 | 0.667 | 0.214 | 0.014 | 0.000 | 0.000 | 0.32 | 0.21 | 0.21 | 0.38 |
| 1998 | 1.694 | 7.692 | 0.569 | 0.120 | 0.000 | 0.000 | 0.000 | 0.21 | 0.16 | 0.30 | 0.53 |
| 1999 | 0.495 | 2.531 | 2.419 | 0.153 | 0.028 | 0.000 | 0.000 | 0.27 | 0.20 | 0.15 | 0.43 |
| 2000 | 6.296 | 1.011 | 0.346 | 0.330 | 0.000 | 0.023 | 0.000 | 0.36 | 0.13 | 0.31 | 0.44 |
| 2001 | 4.067 | 5.614 | 0.184 | 0.058 | 0.040 | 0.000 | 0.000 | 0.29 | 0.15 | 0.39 | 0.42 |
| 2002 | 6.622 | 2.533 | 3.335 | 0.000 | 0.000 | 0.011 | 0.000 | 0.59 | 0.19 | 0.38 | - |
| 2003 | 0.739 | 10.792 | 1.041 | 0.327 | 0.037 | 0.030 | 0.058 | 0.32 | 0.21 | 0.30 | 0.26 |
| 2004 | 2.170 | 1.720 | 0.886 | 0.054 | 0.044 | 0.000 | 0.000 | 0.57 | 0.30 | 0.21 | 0.40 |
| 2005 | 0.635 | 2.251 | 0.294 | 0.280 | 0.183 | 0.000 | 0.000 | 0.56 | 0.29 | 0.60 | 0.64 |
| 2006 | 1.700 | 1.308 | 0.583 | 0.025 | 0.000 | 0.000 | 0.011 | 0.52 | 0.26 | 0.37 | 0.71 |
| 2007 | 1.644 | 1.244 | 0.306 | 0.051 | 0.000 | 0.000 | 0.000 | 0.41 | 0.21 | 0.38 | 0.66 |
| 2008 | 0.407 | 2.172 | 0.130 | 0.052 | 0.042 | 0.010 | 0.000 | 0.46 | 0.32 | 0.39 | 0.66 |
| 2009 | 1.440 | 0.590 | 0.330 | 0.090 | 0.000 | 0.000 | 0.000 | 0.60 | 0.23 | 0.26 | 0.68 |
| 2010 | 10.221 | 2.090 | 0.147 | 0.023 | 0.000 | 0.000 | 0.000 | 0.59 | 0.22 | 0.34 | 0.66 |

NI-GFS October groundfish survey $\quad$ Numbers per 3-miles (approx. 1-h tow) RSE = approximate relative standard error

| Survey | 0-gp | 1-gp | 2-gp | 3-gp | 4-gp | 5-gp | 6-gp | 7+ | RSE(0gp) | RSE(1gp) | RSE(2gp) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.579 | 11.094 | 0.501 | 0.476 | 0.086 | 0.000 | 0.000 | 0.000 | 0.58 | 0.36 | 0.28 |
| 1993 | 7.808 | 5.532 | 1.464 | 0.008 | 0.000 | 0.000 | 0.000 | 0.034 | 0.43 | 0.84 | 0.34 |
| 1994 | 19.962 | 16.725 | 0.254 | 0.104 | 0.000 | 0.000 | 0.000 | 0.000 | 0.28 | 0.43 | 0.42 |
| 1995 | 7.886 | 12.068 | 0.333 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.55 | 0.91 | 0.38 |
| 1996 | 14.813 | 4.866 | 0.501 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.42 | 0.50 | 0.30 |
| 1997 | 4.204 | 13.222 | 0.972 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.45 | 0.41 | 0.40 |
| 1998 | 0.370 | 3.765 | 1.639 | 0.057 | 0.000 | 0.000 | 0.000 | 0.000 | 0.38 | 0.36 | 0.37 |
| 1999 | 20.225 | 0.585 | 0.325 | 0.095 | 0.000 | 0.000 | 0.000 | 0.000 | 0.34 | 0.68 | 0.43 |
| 2000 | 7.242 | 3.016 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.36 | 0.33 | 1.00 |
| 2001 | 8.411 | 5.068 | 1.099 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.45 | 0.35 | 0.35 |
| 2002 | 0.897 | 4.879 | 0.377 | 0.125 | 0.000 | 0.000 | 0.000 | 0.000 | 0.86 | 0.58 | 0.55 |
| 2003 | 2.759 | 1.614 | 0.294 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.48 | 0.66 | 0.63 |
| 2004 | 4.437 | 5.790 | 0.237 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.30 | 0.48 | 0.75 |
| 2005 | 8.245 | 7.061 | 1.077 | 0.173 | 0.029 | 0.000 | 0.000 | 0.000 | 0.52 | 0.89 | 0.62 |
| 2006 | 1.170 | 1.302 | 0.015 | 0.066 | 0.000 | 0.000 | 0.000 | 0.000 | 0.45 | 0.53 | 1.00 |
| 2007 | 0.068 | 0.870 | 0.000 | 0.030 | 0.000 | 0.000 | 0.000 | 0.000 | 0.66 | 0.80 | - |
| 2008 | 0.190 | 0.170 | 0.170 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.57 | 1.00 | 1.00 |
| 2009 | 5.356 | 2.136 | 0.061 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.33 | 0.76 | 1.00 |

Table 6.2.6. continued.

| Irish GFS. <br> October | Irish groundfish survey of the Irish Sea. RV Celtic Explorer | Total nos. per survey |  |  |  |  |  |  |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
|  | $00-\mathrm{gp}$ | 1-gp | 2-gp | 3-gp | 4-gp | 5-gp | 6-gp | 7+ |
| 2003 | 16 | 29 | 31 | 3 | 1 | 0 |  |  |
| 2004 | 23 | 74 | 7 | 2 | 0 |  |  |  |

UK Fishery Science Partnership western Irish Sea pelagic trawl survey (mean nos. per hour) | (revised) |
| :--- |
| Feb-March |

0-gp
0-gp
2004

UK Fishery Science Partnership eastern Irish Sea otter trawl survey (mean nos. per hour) SSB index = kg/hr Feb-March

|  |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0-g \mathrm{p}$ | 1-gp | 2-gp | 3-gp | 4-gp | 5-gp | 6-gp | 7+ | SSB index |
| 2005 | 0.06 | 4.02 | 0.25 | 0.38 | 0.004 | 0.01 | 0 | 5.97 |  |
| 2006 | 0.83 | 0.77 | 0.67 | 0.007 | 0.042 | 0 | 0.001 | 3.31 |  |
| 2007 | 0.59 | 1.43 | 0.09 | 0.08 | 0 | 0 | 0 | 1.77 |  |
| 2008 | 0.01 | 1.80 | 0.32 | 0.02 | 0.03 | 0.003 | 0.01 | 2.60 |  |
| 2009 | 0.50 | 0.36 | 0.21 | 0.09 | 0.01 | 0.004 | 0.00 | 1.58 |  |
| 2010 | 0.98 | 0.65 | 0.03 | 0.04 | 0.01 | 0.000 | 0.00 | 0.84 |  |

ENG BTS-Sept beam trawl survey. No. per 100km

| September <br> Survey | $0-\mathrm{gp}$ |  |
| :--- | :--- | :--- |


| 1991 |  |
| :---: | :---: |
| 1992 |  |
| 1993 | $\mathbf{2 2}$ |
| 1994 | $\mathbf{3 0}$ |
| 1995 | $\mathbf{4 0}$ |
| 1996 | $\mathbf{2 9}$ |
| 1997 | $\mathbf{3 2}$ |
| 1998 | $\mathbf{2}$ |
| 1999 | $\mathbf{4 9}$ |
| 2000 | $\mathbf{3 7}$ |
| 2001 | $\mathbf{2 4}$ |
| 2002 | $\mathbf{7}$ |
| 2003 | $\mathbf{9}$ |
| 2004 | $\mathbf{2 2}$ |
| 2005 | $\mathbf{4 1}$ |
| 2006 | $\mathbf{6}$ |
| 2007 | $\mathbf{4}$ |
| 2008 | $\mathbf{7}$ |

20096

NIMIKNET pelagic 0-gp index
May-June
$\qquad$

Table 6.2.7. B-Adapt model settings for update run in 2010. Same settings as in 2009.

| Setting | Values |
| :--- | :--- |
| Plus group | 5-plus |
| Fbar range | $2-4$ (arithmetic mean) |
| Year range for tuning VPA | 1992 onwards |
| Surveys after final year of catch data used. | Yes; Fmult $=0.75$ for 2010 WGCSE |
| VPA model or cohort analysis used | v (exact) |
| First age with constant catchability | Entered as 0 for all tuning fleets |
| q-plateau | Entered as 3 for all tuning fleets |
| Tapered time weighting applied | No |
| Number of missing catch multipliers | 10 for WGCSE 2010 (bias estimated from 2000 <br> onwards) |
| No. ages for terminal F mean, and scaling <br> factor for mean | ages $=1 ;$ scaling factor = 1.0; arithmetic mean (i.e. <br> Constraint on F or catch? Stiffness weight $(\lambda)$ |
| Prior weighting of fleets | Constrain F; $\lambda=1.0$ |
| Output tables | None |

Table 6.2.8. Selected diagnostics from update B-Adapt (not bootstrap run).

```
Lowestoft VPA Program
26/05/2010 11:04
```

Adapt Analysis
"IRISH SEA COD WGCSE 2010 COMBPLUSGROUP"
CPUE data from file cod7tun.txt
Catch data for 42 years : 1968 to 2009. Ages 0 to $5+$


Time series weights :
Tapered time weighting not applied
Catchability analysis :

| Fleet | PowerQ <br> ages $<x$ | QPlateau <br> ages $>x$ |  |
| :--- | :--- | :--- | :--- |
| NIGFSMAR(1-4gp) |  | 0 | 3 |
| ScoGFS-Q1 Survey (No |  | 0 | 3 |
| NIGFSOCT(0 2-gp) |  | 0 | 3 |
| ENGBTS-Sept | 0 | 3 |  |
| NIMIKNET | 0 | 3 |  |

Catchability independent of stock size for all ages

Bias estimation :
Bias estimated for the final 10 years.
Oldest age F estimates in 1968 to 2010 calculated as 1.000 * the mean $F$ of ages 3-3
Total F penalty applied $\quad$ lambda $=1.000$

Individual fleet weighting not applied

| INITIAL SSQ = | 1866.72552 |
| :--- | ---: |
| PARAMETERS $=$ | 14 |
| OBSERVATIONS $=$ | 207 |
|  |  |
| SSQ $=$ | 96.2695 |
| QSSQ $=$ | 90.25171 |
| CSSQ $=$ | 6.01779 |
| IFAIL $=$ | 0 |
| IFAILCV $=0$ |  |
|  |  |
| Regression weights |  |


| Fishing mortalities <br> Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.141 | 0.239 | 0.15 | 0.205 | 0.157 | 0.126 | 0.043 | 0.149 | 0.135 | 0.114 |
|  | 1 | 1.172 | 0.733 | 1.243 | 1.108 | 0.77 | 0.789 | 1.133 | 0.932 | 1.088 | 0.924 |
|  | 2 | 1.867 | 1.587 | 1.736 | 1.438 | 1.475 | 1.197 | 2.232 | 1.58 | 1.605 | 1.806 |
|  | 4 | 1.867 | 1.587 | 1.736 | 1.438 | 1.475 | 1.197 | 2.232 | 1.58 | 1.605 | 1.806 |

Population numbers (Thousands)

|  | AGE |  | 0 | 1 | 2 | 3 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |

## Table 6.2.8. Continued.

| Estimated population abundance at 1st Jan 2010 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0.00 \mathrm{E}+00$ | $3.06 \mathrm{E}+03$ | $3.43 \mathrm{E}+02$ | $7.30 \mathrm{E}+01$ | $2.65 \mathrm{E}+01$ |
| Taper weighted geometric mean of the VPA populations: |  |  |  |  |  |
|  | 4.21E+03 | $3.46 \mathrm{E}+03$ | $2.39 \mathrm{E}+03$ | $8.86 \mathrm{E}+02$ | $2.34 \mathrm{E}+02$ |
| Standard error of the weighted Log(VPA populations) : |  |  |  |  |  |
|  | 0.8863 | 0.8861 | 0.7892 | 0.8034 | 1.046 |
| Log population residuals (unweighted). |  |  |  |  |  |
| Fleet: $\operatorname{NIGFSMAR}(1-4 \mathrm{gp})$ |  |  |  |  |  |
| Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time |  |  |  |  |  |
| Age | 1 | 2 | 3 | 4 |  |
| Mean Log q | -1.8478 | -1.2984 | -1.5606 | -1.5606 |  |
| S.E(Log q) | 0.554 | 0.3494 | 0.4746 | 0.7574 |  |

Regression statistics

| Age | Slope | t-value |  | Intercept | RSquare | No Pts |  | s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.85 | 1.064 | 2.7 | 0.76 |  | 17 | 0.4674 | -1.85 |
|  | 2 | 1.11 | -0.899 | 0.65 | 0.82 |  | 17 | 0.39009 | -1.3 |
|  | 3 | 0.9 | 0.629 | 2 | 0.74 |  | 17 | 0.43675 | -1.56 |
|  | 4 | 1.25 | -0.714 | 1.06 | 0.36 |  | 16 | 0.92825 | -1.76 |

Fleet : ScoGFS-Q1 Survey (№
Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 1 | 2 | 3 | 4 |
| :--- | ---: | ---: | ---: | ---: |
| Mean Log q | -5.4514 | -3.7801 | -2.8281 | -2.8281 |
| S. E(Log q) | 0.8646 | 0.44 | 0.6657 | 0.9355 |

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time.


Fleet : NIGFSOCT(0 2-gp)
Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 0 |
| :--- | ---: |
| Mean Log q | -1.8198 |
| S.E(Log q$)$ | 0.9899 |

Regression statistics
Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value | Intercept | RSquare | No Pts |  | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0.54 | 3.79 | 4.49 | 0.81 |  | 18 | 0.40086 | -1.82 |

Fleet : ENGBTS-Sept
Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 0 |
| :--- | ---: |
| Mean $\log \mathrm{q}$ | -4.7801 |
| $\mathrm{~S} . \mathrm{E}(\log \mathrm{q})$ | 0.731 |

Regression statistics
Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value | Intercept | RSquare | No Pts |  | Reg s.e | Mean Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0 | 1.25 | -0.899 | 4.06 | 0.46 | 17 | 0.91942 | -4.78 |  |

Table 6.2.8. Continued.

Fleet : NIMIKNET
Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

Age
Mean L
S.E(Log
$\begin{array}{lr}\text { ge } & 0 \\ \text { an } \log q & -5.6152 \\ \log \text { q) } & 1.2769\end{array}$


Parameter
Age

|  | Survivors | s.e log est |
| :--- | ---: | ---: |
| 0 | 3063.51189 | 0.37111 |
| 1 | 343.568 | 0.3601 |
| 2 | 72.98631 | 0.46768 |
| 3 | 26.48076 | 0.50358 |



Year

| Multiplier |  | s.e log est |
| :---: | :---: | :---: |
| 33 | 1.91165 | 0.22951 |
| 34 | 1.86609 | 0.24785 |
| 35 | 2.4543 | 0.23212 |
| 36 | 3.77883 | 0.23988 |
| 37 | 3.28894 | 0.24589 |
| 38 | 2.6592 | 0.24153 |
| 39 | 3.30645 | 0.22181 |
| 40 | 2.60634 | 0.25001 |
| 41 | 2.63492 | 0.25579 |
| 42 | 2.34227 | 0.28337 |


| 0.13772 | 0.01232 | 0.00955 | 0.00584 |
| ---: | ---: | ---: | ---: |
| 0.01232 | 0.1129 | 0.006 | 0.00076 |
| 0.00955 | 0.016 | 0.21873 | -0.00351 |
| 0.00584 | 0.00776 | -0.00351 | 0.25359 |
| 0.00931 | 0.00907 | 0.00868 | 0.00845 |
| 0.0106 | 0.01029 | 0.00984 | 0.00956 |
| 0.01043 | 0.01014 | 0.0947 | 0.00879 |
| 0.00995 | 0.0097 | 0.00931 | 0.00859 |
| 0.00953 | 0.00952 | 0.00934 | 0.00922 |
| 0.0095 | 0.00957 | 0.00923 | 0.00902 |
| 0.00962 | 0.00906 | 0.00835 | 0.00088 |
| 0.01055 | 0.00764 | 0.00457 | 0.00494 |
| 0.01131 | 0.00546 | 0.01507 | -0.00322 |
| 0.01183 | 0.02419 | 0.00003 | 0.00918 |

0.00931
0.00907
0.00868
0.00845
0.0567
0.01689
0.00801
0.00784
0.00903
0.00966
0.00974
0.00948
0.00912

| 0.0106 | 0.01043 |
| ---: | ---: |
| 0.01029 | 0.01014 |
| 0.00994 | 0.0997 |
| 0.00956 | 0.00879 |
| 0.01689 | 0.00881 |
| 0.06143 | 0.01708 |
| 0.01708 | 0.05388 |
| 0.00687 | 0.01988 |
| 0.00799 | 0.00778 |
| 0.01017 | 0.00977 |
| 0.01105 | 0.00971 |
| 0.01088 | 0.01038 |
| 0.01041 | 0.0101 |


|  |
| :---: |
|  |  |

0.00953
0.00952
0.00934
0.00922
0.00903
0.00799
0.00778
0.0168
0.06046
0.01916
0.00621
0.00648
0.00824
0.0095
0.00957
0.00923
0.0902
0.00966
0.01017
0.00797
0.0073
0.01916
0.05833
0.01532
0.00585
0.0611
0.00962
0.00906
0.00835
0.00788
0.00974
0.01105
0.00971
0.00723
0.00621
0.01532
0.0492
0.01752
0.00797
0.00665
0.02419

00003

Table 6.2.9. Cod in VIIa. Point estimates of population numbers-at-age from the update B-Adapt assessment.

| YEAR | 0 | 1 | 2 | 3 | 4 | 5+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 6512 | 3424 | 3710 | 1600 | 727 | 330 |
| 1969 | 8506 | 5332 | 2475 | 1640 | 420 | 467 |
| 1970 | 15131 | 6964 | 3571 | 711 | 412 | 473 |
| 1971 | 5239 | 12388 | 4516 | 1684 | 268 | 220 |
| 1972 | 13883 | 4289 | 7680 | 1891 | 574 | 269 |
| 1973 | 3107 | 11366 | 2802 | 3367 | 812 | 480 |
| 1974 | 11055 | 2544 | 7270 | 1317 | 1168 | 485 |
| 1975 | 3533 | 9051 | 1606 | 2777 | 580 | 722 |
| 1976 | 5103 | 2893 | 5881 | 740 | 1020 | 336 |
| 1977 | 5529 | 4178 | 1353 | 2135 | 282 | 500 |
| 1978 | 12082 | 4527 | 2686 | 650 | 652 | 218 |
| 1979 | 14196 | 9892 | 3087 | 1222 | 255 | 472 |
| 1980 | 7923 | 11623 | 6513 | 1376 | 459 | 196 |
| 1981 | 3461 | 6487 | 7238 | 2832 | 477 | 286 |
| 1982 | 5264 | 2833 | 4142 | 2685 | 1028 | 334 |
| 1983 | 7879 | 4310 | 2009 | 1359 | 903 | 362 |
| 1984 | 7922 | 6451 | 2796 | 813 | 444 | 437 |
| 1985 | 6350 | 6486 | 3864 | 1221 | 274 | 280 |
| 1986 | 18442 | 5199 | 4214 | 1290 | 375 | 238 |
| 1987 | 8743 | 15099 | 3380 | 1448 | 434 | 203 |
| 1988 | 3803 | 7158 | 8481 | 1170 | 438 | 182 |
| 1989 | 4904 | 3113 | 3361 | 2732 | 335 | 178 |
| 1990 | 5648 | 4015 | 2025 | 835 | 570 | 153 |
| 1991 | 8751 | 4624 | 2648 | 701 | 201 | 190 |
| 1992 | 1709 | 7165 | 2022 | 914 | 182 | 125 |
| 1993 | 5110 | 1399 | 4629 | 553 | 163 | 40 |
| 1994 | 3699 | 4184 | 945 | 1212 | 97 | 43 |
| 1995 | 3121 | 3028 | 2751 | 268 | 243 | 19 |
| 1996 | 5793 | 2555 | 2031 | 1107 | 61 | 71 |
| 1997 | 2105 | 4743 | 1806 | 672 | 285 | 44 |
| 1998 | 881 | 1723 | 3411 | 460 | 110 | 49 |
| 1999 | 5656 | 721 | 1227 | 1080 | 81 | 34 |
| 2000 | 3985 | 4630 | 528 | 259 | 121 | 13 |
| 2001 | 4652 | 3263 | 3293 | 134 | 33 | 29 |
| 2002 | 1234 | 3808 | 2104 | 1295 | 22 | 16 |
| 2003 | 2074 | 1010 | 2685 | 497 | 187 | 6 |
| 2004 | 1269 | 1698 | 674 | 726 | 97 | 39 |
| 2005 | 1491 | 1039 | 1189 | 256 | 136 | 24 |
| 2006 | 1236 | 1221 | 750 | 442 | 63 | 46 |
| 2007 | 384 | 1012 | 958 | 198 | 39 | 3 |
| 2008 | 574 | 314 | 714 | 309 | 33 | 10 |
| 2009 | 3742 | 470 | 225 | 197 | 51 | 10 |
| 2010 | 0 | 3064 | 343 | 73 | 26 | 8 |

Table 6.2.10. Cod in VIIa. Point estimates of fishing mortality-at-age from the update B-Adapt assessment. Figures for 2010 are the values assumed for a $25 \%$ reduction in F in the intermediate year.


Table 6.2.11. Cod in VIIa. Summary data from the update B-Adapt assessment. "B-Adapt removals" are the estimated total removals from 2000 onwards in excess of removals due to the assumed natural mortality rate.

Summary (without SOP correction)

| Year | $\begin{gathered} \text { Recruits age } 0 \\ \text { (thousands) } \\ \hline \end{gathered}$ | Total biomass (t) | $\begin{gathered} \text { Spawning stock } \\ \text { biomass (t) } \\ \hline \end{gathered}$ | $\begin{aligned} & \text { Input landings } \\ & (\mathrm{t}) \\ & \hline \end{aligned}$ | $\begin{gathered} \text { B-Adapt } \\ \text { removals ( } \mathrm{t} \text { ) } \end{gathered}$ | FBAR 2-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 6512 | 19351 | 13444 | 8541 |  | 0.9634 |
| 1969 | 8506 | 18040 | 12241 | 7991 |  | 1.1365 |
| 1970 | 15131 | 17709 | 9785 | 6426 |  | 0.7005 |
| 1971 | 5239 | 23476 | 11271 | 9246 |  | 0.808 |
| 1972 | 13883 | 26393 | 15873 | 9234 |  | 0.6386 |
| 1973 | 3107 | 30044 | 20227 | 11819 |  | 0.7572 |
| 1974 | 11055 | 27155 | 18121 | 10251 |  | 0.6671 |
| 1975 | 3533 | 25060 | 17886 | 9863 |  | 0.7262 |
| 1976 | 5103 | 21465 | 13647 | 10247 |  | 0.7806 |
| 1977 | 5529 | 16614 | 12673 | 8054 |  | 0.8352 |
| 1978 | 12082 | 14188 | 8662 | 6271 |  | 0.6851 |
| 1979 | 14196 | 19638 | 10426 | 8371 |  | 0.723 |
| 1980 | 7923 | 26103 | 12310 | 10776 |  | 0.7836 |
| 1981 | 3461 | 29723 | 18317 | 14907 |  | 0.8062 |
| 1982 | 5264 | 27025 | 20249 | 13381 |  | 0.8981 |
| 1983 | 7879 | 21842 | 15260 | 10015 |  | 0.8471 |
| 1984 | 7922 | 18773 | 11249 | 8383 |  | 0.7999 |
| 1985 | 6350 | 21980 | 12055 | 10483 |  | 0.9532 |
| 1986 | 18442 | 20979 | 12026 | 9852 |  | 0.8823 |
| 1987 | 8743 | 28289 | 12995 | 12894 |  | 0.9503 |
| 1988 | 3803 | 26056 | 13492 | 14168 |  | 1.0121 |
| 1989 | 4904 | 21061 | 14300 | 12751 |  | 1.3086 |
| 1990 | 5648 | 14540 | 8725 | 7379 |  | 1.1025 |
| 1991 | 8751 | 13177 | 6531 | 7095 |  | 1.0541 |
| 1992 | 1709 | 15518 | 7231 | 7735 |  | 1.3814 |
| 1993 | 5110 | 12376 | 6295 | 7555 |  | 1.4055 |
| 1994 | 3699 | 10460 | 5995 | 5402 |  | 1.2903 |
| 1995 | 3121 | 10439 | 4575 | 4587 |  | 1.0964 |
| 1996 | 5793 | 10298 | 5747 | 4964 |  | 1.072 |
| 1997 | 2105 | 11795 | 5614 | 5859 |  | 1.4641 |
| 1998 | 881 | 9888 | 4810 | 5318 |  | 1.338 |
| 1999 | 5656 | 6769 | 4918 | 4784 |  | 1.7764 |
| 2000 | 3985 | 6630 | 2040 | 1274 | 1.912 | 1.6356 |
| 2001 | 4652 | 10194 | 3242 | 2252 | 1.866 | 1.3025 |
| 2002 | 1234 | 12179 | 6197 | 2695 | 2.454 | 1.5715 |
| 2003 | 2074 | 8389 | 4405 | 1285 | 3.779 | 1.3282 |
| 2004 | 1269 | 6949 | 4140 | 1072 | 3.289 | 1.2398 |
| 2005 | 1491 | 5067 | 2690 | 910 | 2.659 | 1.0612 |
| 2006 | 1236 | 4623 | 2757 | 840 | 3.306 | 1.8661 |
| 2007 | 384 | 3600 | 1658 | 702 | 2.606 | 1.3642 |
| 2008 | 574 | 2784 | 1801 | 662 | 2.635 | 1.4327 |
| 2009 | 3742 | 1987 | 1192 | 468 | 2.342 | 1.5121 |
| $\begin{aligned} & \hline \text { Average } \\ & \text { (1968-2009) } \end{aligned}$ | 5754 | 16158 | 9454 | 7066 | 2.685 | 1.0942 |

Table 6.2.12. Cod VIIa : Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships and the yield-per-recruit $F_{\text {msy }}$ proxies.

Stock name
Cod VIIa
Sen filename
codviia.sen
pf, pm
$0 \quad 0$
Number of iterations 1000
Simulate variation in Biological parameters TRUE
SR relationship constrained
TRUE
Ricker
971/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 1.82 | 0.57 | 22731 | 11242 | 0.78 | 0.69 | 0.94 | 0.00 | 78.86 |
| Mean | 1.88 | 0.56 | 32311 | 14620 | 0.80 | 0.67 | 0.96 | 0.00 |  |
| 5\%ile | 0.91 | 0.35 | 14354 | 7624 | 0.64 | 0.22 | 0.64 | 0.00 |  |
| 25\%ile | 1.25 | 0.45 | 18967 | 9788 | 0.72 | 0.48 | 0.80 | 0.00 |  |
| 50\%ile | 1.63 | 0.54 | 23850 | 11721 | 0.79 | 0.66 | 0.92 | 0.00 |  |
| 75\%ile | 2.26 | 0.65 | 31879 | 14553 | 0.86 | 0.85 | 1.10 | 0.00 |  |
| 95\%ile | 3.73 | 0.84 | 66037 | 26736 | 0.98 | 1.14 | 1.40 | 0.00 |  |
| CV | 0.46 | 0.27 | 1.32 | 1.10 | 0.13 | 0.41 | 0.24 | 0.41 |  |

Beverton-Holt
922/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 1.93 | 0.27 | 67633 | 17521 | 0.67 | 1.28 | 12077.10 | 12108.80 | 79.72 |
| Mean | 1.93 | 0.24 | 169949 | 29096 | 0.68 | 1.33 | 20432.23 | 28075.58 |  |
| 5\%ile | 0.88 | 0.12 | 31945 | 9301 | 0.23 | 1.07 | 6961.25 | 4039.37 |  |
| 25\%ile | 1.26 | 0.20 | 50861 | 13369 | 0.47 | 1.21 | 9205.57 | 7871.53 |  |
| 50\%ile | 1.68 | 0.25 | 74698 | 17667 | 0.68 | 1.32 | 11891.70 | 12532.45 |  |
| 75\%ile | 2.37 | 0.29 | 122226 | 26686 | 0.88 | 1.44 | 17385.50 | 22701.85 |  |
| 95\%ile | 3.87 | 0.36 | 354280 | 54120 | 1.17 | 1.63 | 36099.99 | 60958.19 |  |
| CV | 0.48 | 0.31 | 4.85 | 5.25 | 0.43 | 0.13 | 5.42 | 7.09 |  |

Smooth hockeystick
997/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 1.29 | 0.38 | 33581 | 11734 | 0.54 | 0.99 | 0.33 | 10186.80 | 76.49 |
| Mean | 1.36 | 0.36 | 55050 | 12815 | 0.56 | 1.04 | 0.34 | 10624.59 |  |
| 5\%ile | 0.74 | 0.14 | 20209 | 8067 | 0.44 | 0.76 | 0.27 | 7754.99 |  |
| 25\%ile | 0.97 | 0.28 | 28227 | 10494 | 0.50 | 0.90 | 0.30 | 9241.65 |  |
| 50\%ile | 1.21 | 0.36 | 35922 | 12276 | 0.55 | 1.01 | 0.33 | 10404.80 |  |
| 75\%ile | 1.59 | 0.43 | 47836 | 14611 | 0.60 | 1.13 | 0.37 | 11630.70 |  |
| 95\%ile | 2.47 | 0.58 | 99557 | 19006 | 0.69 | 1.42 | 0.42 | 14588.56 |  |
| CV | 0.44 | 0.36 | 1.76 | 0.28 | 0.14 | 0.20 | 0.14 | 0.20 |  |
| Per recruit |  |  |  |  |  |  |  |  |  |
|  | F35 | F40 | F01 | Fmax | Bmsypr | MSYpr | Fpa | Flim |  |
| Deterministic | 0.24 | 0.20 | 0.19 | 0.38 | 5.01 | 1.75 | 0.4 | 0.72 |  |
| Mean | 0.23 | 0.19 | 0.18 | 0.36 | 7.85 | 1.82 |  |  |  |
| 5\%ile | 0.05 | 0.04 | 0.05 | 0.14 | 3.04 | 1.27 |  |  |  |
| 25\%ile | 0.16 | 0.13 | 0.14 | 0.28 | 4.16 | 1.54 |  |  |  |
| 50\%ile | 0.23 | 0.19 | 0.18 | 0.36 | 5.18 | 1.75 |  |  |  |
| 75\%ile | 0.29 | 0.24 | 0.22 | 0.43 | 6.73 | 2.05 |  |  |  |
| 95\%ile | 0.38 | 0.32 | 0.29 | 0.58 | 13.87 | 2.59 |  |  |  |
| CV | 0.43 | 0.43 | 0.39 | 0.36 | 1.72 | 0.23 |  |  |  |



Figure 6.2.1. Cod in VIIa. Catch weights-at-age (same as stock weights).


Figure 6.2.2. Cod in VIIa. Landings number-per-age.


Figure 6.2.3. Cod in VIIa. Landings-per-age as 3D bars.


Figure 6.2.3b. Cod in VIIa. Landings-per-age as 3D bars alternative perspective.


Figure 6.2.4. Cod in VIIa. Landings data used in the B-Adapt assessment.


Figure 6.2.5. Cod in VIIa. Separable VPA residuals.









Number of trips sampled

|  | Demersal | Nephrops | Beam | All |
| :---: | :---: | :---: | :---: | :---: |
| 2004 | 7 | 7 | 0 | 14 |
| 2005 | 4 | 2 | 2 | 8 |
| 2006 | 3 | 1 | 1 | 5 |
| 2007 |  |  |  | 26 |
| 2008 |  |  |  | 29 |
| 2009 |  |  |  | 11 |

Figure 6.2.6. Cod in VIIa. Length frequencies of retained and discarded cod recorded by observers on UK (E\&W) fishing vessels in 2004-2009 (nos. for observed trips).


Figure 6.2.7. Cod in VIIa. Length frequencies of retained and discarded cod recorded by observers on Irish otter trawl vessels in 2009, raised to fleet level (no. trips sampled =12).


Figure 6.2.8. Cod in VIIa. Log survey indices for NIGFS-Mar and ScoGFS-Q1 trawl surveys by year and year class; comparative scatterplots of indices within year classes, and residuals from Surba model fits.


Figure 6.2.9. Cod in VIIa. Consistency between trends in year-class strength estimated from SURBA analysis of NIGFS-Mar and ScoGFS-Q1 surveys and the other 0-gp indices used in the assessment.


Figure 6.2.10. Trends in empirical SSB indices from 2004 onwards from the NIGFS-Mar compared with equivalent indices from UK Fisheries Science Partnership surveys of the western and eastern Irish Sea in February-March.


Figure 6.2.11. Cod in VIIa: Catchability residuals from the update B-Adapt run (non-bootstrap option).


Figure 6.2.12. Comparison plots for non-bootstrap B-Adapt cod assessments with and without the bias estimated.


Figure 6.2.13. Retrospective plots for B-Adapt cod assessment. All runs use the non-bootstrap option and therefore give point estimates rather than bootstrap 50th percentiles.


Figure 6.2.14. Stock summary plot from update B-Adapt run. Continuous line on landings plot is the reported landings; filled squares are landings in 1991-2002 and 2005 including sample-based estimates at three ports; open circles with $90 \%$ confidence intervals are total removals estimates (in excess of assumed natural mortality) from B-Adapt. Dotted lines on plots are 5th and 95th bootstrap percentiles.


Figure 6.2.15. Cod in VIIa: comparison of updated B-ADAPT stock trends with indices of recruitment, SSB and fishing mortality from SURBA runs with NIGFS-Mar and ScoGFS-Q1 surveys. The $B$-Adapt estimates of $F$ have been increased by $M=0.2$ to give $Z$ indices comparable with the SURBA values.


Figure 6.2.16. Cod in VIIa. Stock-recruit data with segmented regression model fitted assuming log-normal variability in recruitment. The most recent 7 year classes are indicated by open symbols.

| Fbar(2-4) | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 0.86 | 0.00 | 0.00 |
| 0.25 | 1.00 | 0.00 | 0.00 |
| 0.5 | 1.14 | 0.00 | 0.00 |
| 0.75 | 1.28 | 0.00 | 0.00 |
| 0.95 | 1.54 | 0.00 | 0.00 |


Zero F projection
$25 \%$ F reduction in 2010
Recruitment 2002-2008

Figure 6.2.17. Cod in VIIa. Projection to 2015 based on the update B-Adapt assessment, assuming $25 \%$ F reduction in 2010 and zero F in subsequent years. Recruitment is bootstrapped from the 2002-2008 year classes. Percentiles of F, SSB and removals, and probability of SSB>Blim, are tabulated for selected years.

| Fbar(2-4) | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 0.86 | 0.00 | 0.00 |
| 0.25 | 1.00 | 0.00 | 0.00 |
| 0.5 | 1.14 | 0.00 | 0.00 |
| 0.75 | 1.28 | 0.00 | 0.00 |
| 0.95 | 1.54 | 0.00 | 0.00 |




| Removals | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 660 | 0 | 0 |
| 0.25 | 812 | 0 | 0 |
| 0.5 | 934 | 0 | 0 |
| 0.75 | 1071 | 0 | 0 |
| 0.95 | 1365 | 0 | 0 |




Figure 6.2.18. Cod in VIIa. Projection to 2015 based on the update B-Adapt assessment, assuming $25 \%$ F reduction in 2010 and zero F in subsequent years. Recruitment is bootstrapped from the 2002-2009 year classes. Percentiles of F, SSB and removals, and probability of SSB>Blim, are tabulated for selected years.

| Fbar(2-4) | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 0.86 | 0.64 | 0.48 |
| 0.25 | 1.00 | 0.75 | 0.56 |
| 0.5 | 1.14 | 0.86 | 0.64 |
| 0.75 | 1.28 | 0.96 | 0.72 |
| 0.95 | 1.54 | 1.16 | 0.87 |
| SSB | Year |  |  |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 500 | 1258 | 2206 |
| 0.25 | 631 | 1822 | 3552 |
| 0.5 | 744 | 2249 | 4626 |
| 0.75 | 873 | 2873 | 6049 |
| 0.95 | 1082 | 4027 | 8727 |
| Removals | Year |  |  |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 660 | 1124 | 1283 |
| 0.25 | 812 | 1647 | 1892 |
| 0.5 | 934 | 2147 | 2442 |
| 0.75 | 1071 | 2737 | 3120 |
| 0.95 | 1365 | 4069 | 4412 |



Figure 6.2.19. Cod in VIIa. Projection to 2015 based on the update B-Adapt assessment, assuming $25 \%$ annual $F$ reduction in 2010 until the year when median $F$ reaches a value of 0.4 . Recruitment is bootstrapped from the 2002-2008 year classes. Percentiles of F, SSB and removals, and probability of SSB>Blim, are tabulated for selected years.

| Fbar(2-4) | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 0.86 | 0.64 | 0.48 |
| 0.25 | 1.00 | 0.75 | 0.56 |
| 0.5 | 1.14 | 0.86 | 0.64 |
| 0.75 | 1.28 | 0.96 | 0.72 |
| 0.95 | 1.54 | 1.16 | 0.87 |



| Removals | Year |  |  |
| :---: | :---: | :---: | :---: |
| Percentile | 2010 | 2011 | 2012 |
| 0.05 | 660 | 1137 | 1360 |
| 0.25 | 812 | 1666 | 1969 |
| 0.5 | 934 | 2154 | 2565 |
| 0.75 | 1071 | 2764 | 3332 |
| 0.95 | 1365 | 4100 | 4897 |

Figure 6.2.20. Cod in VIIa. Projection to 2015 based on the update B-Adapt assessment, assuming $25 \%$ annual $F$ reduction in 2010 until the year when median $F$ reaches a value of 0.4 . Recruitment is bootstrapped from the 2002-2009 year classes. Percentiles of F, SSB and removals, and probability of SSB>Blim, are tabulated for selected years.


Figure 6.2.21. Cod in VIIa. Stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $F_{\text {msy. }}$. Right hand panels : curves plotted from the first 100 MCMC re-samples with converged $\mathrm{F}_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of Fmš from the $\mathbf{1 0 0 0}$ re-samples.

## Cod VIla Beverton-Holt



Figure 6.2.22. Cod in VIIa. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Beverton and Holt stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy. }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

## Cod VIla Smooth hockeystick



Figure 6.2.23. Cod in VIIa. Estimates of $F$ reference points and equilibrium yield and SSB against fishing mortality using a Hockeystick stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

## Cod VIla Ricker



Figure 6.2.24. Cod in VIIa. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Ricker stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $F_{m s y}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.


Figure 6.2.25. Cod in VIIa. Fitted yield-per-recruit F reference points, yield-per-recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric resampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first 100 re-samples.


Figure 6.2.26. Cod in VIIa. Annual Egg Production Method (AEPM) distribution of Stage 1 cod eggs during 2008. Station positions are marked with crosses, and the stratum boundaries are indicated.


Figure 6.2.27. Cod in VIIa. Time-series of Annual Egg Production Method (AEPM) estimates of SSB ( +2 SE) relative to ICES estimates (ICES, 2009). "B-Adapt SSB" is the series given by ICES (2009). "Variable maturity SSB" is the ICES estimates adjusted to reflect a time-series of maturity from the AFBI groundfish surveys (see WD11).

### 6.3 Haddock in Division VIIa

## Type of assessment

The Working Group performed an update assessment for this stock in 2010.

## ICES advice applicable to 2009

The advice from ICES for 2009, under single-stock exploitation boundaries, was as follows:

Exploitation boundaries in relation to precautionary limits: The available information is inadequate to evaluate spawning-stock or fishing mortality relative to precautionary reference points. SSB is increasing and recent recruitments appear to be above average. ICES recommends that fishing effort should not be allowed to increase.

## ICES advice applicable to 2010

The state of the stock is uncertain. Stock trends indicate an increase in SSB over the time-series but a decrease in 2008. Recruitment in the last two years appears to be below average. Total mortality appears relatively stable. ICES advises on the basis of precautionary considerations that there should be no increase in effort relative to 2009.

### 6.3.1 General

## Stock descriptions and management units

The stock and management units are both ICES Division VIIa (Irish Sea).

## Management applicable to 2009 and 2010

Management measures include TAC and effort restrictions as well as technical measures. Due to the bycatch of cod in the haddock fishery, the regulations affecting Irish Sea haddock remain linked to those implemented under the cod recovery plan.

TAC regulations for 2009 and 2010 are given below:

| Species: Haddock <br> Melanognammus aeglefinus | Zone:VIIa <br> (HAD/07A.) |  |
| :--- | ---: | ---: | :--- |
| Belgium | 23 |  |
| France | 103 |  |
| Ireland | 617 |  |
| United Kingdom | 681 |  |
| EC | 1424 |  |
| TAC | 1424 | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

2010

| Species: | Haddock <br> Melanogrammus acglefinus | Zone:VIla <br> (HAD/07A) |
| :--- | :--- | :--- | :--- |
| Belgium | 23 |  |
| France | 103 |  |
| Ireland | 617 |  |
| United Kingdom | 681 |  |
| EU | 1424 | Precautionary TAC |
| TAC | 1424 |  |

The minimum landing size for haddock in the Irish Sea is 30 cm .

## Fishery in 2009

The characteristics of the fishery are described in the Stock Annex. An overview of the fisheries in the Irish Sea is given in Section 6.1.

The fishery in 2009 was prosecuted by the same fleets and gears as in recent years, with directed fishing prevented inside the cod closure in spring. The targeted whitefish fishery that developed during the 1990 using semi-pelagic trawls, continued to decline during 2009.

The reported uptake of TAC has been poor since 2004, with the exception of 2007. The estimated percentage uptake of UK, Irish and Belgium vessels in 2009 were $68 \%$ (estimated 460 t of 681 t quota), $60 \%$ ( 372 t of 617 t ) and $29 \%$ ( 7 t of 23 t ), respectively. For these figures, quota swaps have, however, not been taken into account. No French landings were reported to the Working Group.

Table 6.3.1 gives nominal landings of haddock from the Irish Sea (Division VIIa) as reported by each country to ICES since 1984.

### 6.3.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1. The landings of the fleets sampled by quarter comprise $70 \%$ of the international total in
2009. No sampling information is available for some of the smaller fleets contributing to the international landings.

## Landings

Table 6.3.2 gives the long-term trend of nominal landings of haddock from the Irish Sea (Division VIIa) as reported to ICES since 1972, together with Working Group estimates. The 1993-2005 WG estimates (excl. 2003) include sampled-based estimates of landings into a number of Irish Sea ports. Sampled-based evidence suggests that WG estimates are close to reported landings since 2006.

The methods for estimating quantities and composition of haddock landings from VIIa, used in previous years, are described in the Stock Annex (Annex 6.3). The series of numbers-at-age in the international commercial landings is given in Table 6.3.3. Sampling levels were not considered adequate to derive catch age compositions in 2003. The time-series mean weight-at-age in the landings is given Table 6.3.4.

## Discards

The series of the Irish and Northern Irish discard data, raised to the number of trips, were updated. Discard numbers-at-age for the different sampled fleets are given in Table 6.3.5. The proportions of discards-by-age for the different sampled fleets are given in Table 6.3.6. There are various issues relating to the reliability of the data, which needs to be addressed at the next benchmark assessment for this stock.

Methods for estimating quantities and composition of discards from UK (NI) and Irish Nephrops trawlers are described in the Stock Annex (Annex 6.3). Sampling levels have increased in recent years, but the highly variable. The very large estimates of discarding for Nephrops fleets observed by previous WG are still evident.

## Biological data

The derivation of biological parameters and variables is described in the Stock Annex Natural mortality was assumed as 0.2 for all ages and years, and proportion mature knife-edged at age 2 for all years.

There is evidence for a decline in mean length of adult haddock over time (Figure 6.3.1), which needs to be reflected in the stock weights-at-age. Since 2001 the WG calculated stock weights by fitting a von Bertalanffy growth curve to all available survey estimates of mean length-at-age in March, described in the Stock Annex 6.3. The procedure was updated this year using NIGFS-Mar and quarter one commercial landings data for 2008. The time-series of length-weight parameters indicate a reduction in expected weight-at-length since 1996 (see Stock Annex for historical data):

| Length-weight parameters |  |  |  | Expected weight-at-length |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | A | B | 30 cm |  | 40 cm |
| 2005 | 0.00489 | 3.174 | 238 | 593 |  |
| 2006 | 0.00506 | 3.165 | 239 | 595 |  |
| 2007 | 0.00469 | 3.194 | 244 | 612 |  |
| 2008 | 0.00523 | 3.159 | 242 | 601 |  |
| 2009 | 0.00431 | 3.224 | 249 | 629 |  |
| 2010 | 0.00413 | 3.238 | 250 | 635 |  |

The following parameter estimates were obtained (last year's estimates in parentheses):

$$
\text { Mean } \mathrm{LI}_{\mathrm{yc}}=80.3 \mathrm{~cm}(80.5) ; \mathrm{K}=0.191(0.191) ; \mathrm{t}_{0}=-0.418(-0.419)
$$

Year-class effects giving estimates of asymptotic length relative to the mean were as follows (2008 and 2009 data were combined as there is only one observation for the 2009 year-class):

| Year class | Effect |  | Year class |
| :---: | :---: | :---: | :---: |
| 1990 | 1.230 | 2000 | 0.973 |
| 1991 | 1.168 | 2001 | 1.000 |
| 1992 | 1.099 | 2002 | 0.962 |
| 1993 | 1.114 | 2003 | 0.905 |
| 1994 | 1.128 | 2004 | 0.834 |
| 1995 | 1.100 | 2005 | 0.858 |
| 1996 | 1.012 | 2006 | 0.854 |
| 1997 | 0.988 | 2007 | 0.898 |
| 1998 | 1.000 | $2008 / 2009$ | 0.923 |
| 1999 | 0.954 |  |  |

The year-class effects show a smooth decline from the mid-1990s coincident with the rapid growth of the stock and may represent density-dependent growth effects, although other environmental factors may contribute. The close fit of the model to observed length-at-age data is shown by year class in Figure 6.3.1. The resultant stock weights-at-age are given in Table 6.3.7.

## Surveys

The survey data considered in the assessment for this stock are given in Table 6.3.8. Survey-series for haddock available to the Working Group are described in the Stock Annex for 7a haddock. The following age-structured abundance indices were used in the assessment:

- UK (NI) groundfish survey (NIGFS) in March (age classes 1 to 5, years 1992-2010).

Additional age-structured abundance indices, that provided auxiliary information, are available from the following sources:

- UK (NI) groundfish survey (NIGFS) in October (age classes 0 to 3; years 1991 to 2009).
- UK (NI) Methot-Isaacs Kidd (MIK) net survey in June (age 0; years 19942009).
- UK Fishery Science Partnership (FSP) Irish Sea roundfish survey, 200420010(www.cefas.co.uk/fsp)
- UK Irish Sea Annual Egg Production Method survey (AEPM), 2006-2008 (Armstrong et al., WD11).

The relative abundance indices are plotted against time in Figure 6.3.2. Surveys give similar signals for all ages ( $0-4$ ). The two 0 -group indices indicate increased recruitment in 2009 after two years of below average recruitment. Strong year classes were evident for all age groups in all surveys, indicating that the different surveys were
capturing the prominent year-class signals in this stock (Figure 6.3.3). Correlation between survey indices by age is positive for all surveys and show high consistency within each fleet, but patchy consistency between the fleets (Stock Annex 6.3). The indices from the UK FSP survey ((Armstrong et al., WD10) in the western Irish Sea also show similar year class signals to the other survey-series, but are noisy with obvious year effects (Figure 6.3.2). Haddock SSB estimates derived from an annual egg production method in the Irish Sea show a similar increase from 2006-2008 as the SURBA estimates from NIGFS-Mar data (Figure 6.3.4). The international landings-atage (excl. 2003) show similar patterns of year-class variation to the surveys (Figure 6.3.2), giving confidence in the combined ability of the surveys to track year classes through time. The signal from the landings-at-age data is, however, much reduced since 2004.

The empirical trend in SSB from both the NIGFS series show the growth in SSB in the mid-1990s, a decline to 2000 and a subsequent variable trend (Figure 6.3.5). In recent years, both surveys show a marked increasing trend in SSB from 2005-2007 and then a decreasing trend to 2009 (diverging considerably in 2008).

## Commercial cpue

Commercial cpue data are available for this stock but are not currently used in the assessment.

## Other relevant data

An IBTS-coordinated UK trawl survey started in the Irish Sea in November/December 2004. Survey index data from this survey have not yet been provided to the Working Group.

### 6.3.3 Historical stock development

## Deviation from Stock Annex

The assessment presented is the single fleet SURBA analysis, using only the NIGFSMar survey. The assessment does not deviate from the procedure used last year, as described in the Stock Annex.

SURBA 3.0 was used for the assessment and model settings (similar to last year's assessment) are given below:

|  | WGCSE 2010 |  |
| :--- | :---: | :---: |
| Year range: |  | 1992-2010 |
| Age range: |  | $1-5$ |
| Catchability: | 1.0 at all ages |  |
| Age weighting | 1.0 at all ages |  |
| Smoothing (Lambda): |  | 1.0 |
| Cohort weighting: | not applied |  |
| Reference age |  | 2 |
| Survey used | NIGFS-Mar |  |

## Data screening

Screening of internal and between survey consistency is described in Section 6.3.2.

## Final update assessment

SURBA model residuals (log-population indices) for the NIGFS-Mar survey show noisy residuals (Figure 6.3.6). Residuals show some evidence of year effects in older ages in some years. The age 2 residual pattern from the NIGFS-Mar survey continue to show a better pattern than the other ages. The NIGFS-Mar survey model show quite large retrospective patterns in SSB (Figure 6.3.6) during the early 2000s, probably related to an overestimation of the 2001 year class. There are also large retrospective patterns in mortality estimates, highlighting the difficulty in estimating mortality for this stock.

The trends in Z, SSB and recruitment for the assessment using the NIGFS-Mar survey data, and the model residuals are given in Figures 6.3.7 and 6.3.8. The SURBA fitted numbers-at-age and total mortality-at-age given in Table 6.3.9. The SURBA index of $Z$ generally follows the much noisier empirical estimates. Both the empirical and SURBA estimates of SSB give a similar increasing trend from 2005-2008 followed by in decrease since 2009. The recruitment estimates at age 1 indicate an above average recruitment in 2009, following two years of poorer recruitment. In general, the SURBA results capture similar year-class dynamics than observed from the raw survey indices (Figure 6.3.2).

## Comparison with previous assessments

The perception of the stock has not changed since last year's assessment. Figure 6.3.9 compares the relative trends between the SURBA fitted estimates from this year's to last year's assessment. The two series show similar trends. The most recent SSB estimate indicates that the stock has decline further since last year. The relative SSB estimate for 2010 is below the series average.

## State of the stock

Stock trends indicate an increase in SSB over the time-series. SSB trend is declining since 2008. The stock is characterised by highly variable recruitment. The model indicates above average recruitment for the 2009 year class after below average recruitment for the 2007 and 2008 year classes. Total mortality remains stable.

### 6.3.4 Short-term projections

No short-term forecast has been performed for this stock. This year the WG projected the SSB for 2011 using the 2010 survey information. Since maturity for the stock is considered as knife-edge at age 2, all the age classes that will comprise the 2011 SSB are already represented by the 2010 quarter one survey index. SSB for 2011 was projected using an average of the last three years total mortality from the SURBA model, a three year average of stock weights (2008-2010) and 10-year geometric mean recruitment.

The projected SSB trend is illustrated in Figure 6.3.10, indicating a stabilisation of the decreasing trend in SSB. SURBA fitted recruitment estimates are also compared to recruitment from the 0-gp indices (NIGF-Oct and NIMIK), indicating that the model estimates might overestimate the strength of the 2007 and 2008 year classes, suggesting that the projected SSB might also be an overestimate.

### 6.3.5 MSY evaluations

MSY evaluations were performed on a very limited dataset. Input data were taken from the last accepted catch-at-age assessment in 2002 from the ICES network (similar
input data to the yield-per-recruit analysis presented in Table 6.3.11). The analysis was performed using the srmsymc ADMB package. Recent assessments were based on survey data only due to the uncertainty with catch-at-age data. This evaluation is based on this historical catch-at-age data, including the underlying problems with the accuracy of the data.

The three stock-recruit relationships fitted by srmsymc are illustrated in Figure 6.3.11. The high uncertainty around these fits reflects the shortage of information within the limited dataseries to inform any stock-recruit relationship. The data are very noisy with relatively high rejection rates for the Ricker and Beverton-Holt models. Mathematically there is very little to distinguish between the three models, based on the AIC values that indicate equal fits (Table 6.3.10). F reference points are poorly defined with wide distributions and very high levels of uncertainty (cv values are high for all three models). $\mathrm{F}_{\text {msy }}$ values falls within the range of $\mathrm{F}_{\text {crash }}$ in all cases (Table 6.3.10).

Stock-recruit relationships are generally poorly defined for haddock stocks. These models assume a positive relationship between spawning-stock size and recruitment. However, haddock is characterised by sporadic high recruitment even at low spawn-ing-stock levels making any relationship difficult to define. Recent trends within the Irish Sea haddock stock showed that an increase in spawning-stock biomass is dependent on these impulses of high recruitment, i.e. recruit-stock. Density-dependent growth is also evident by year class, which will have an effect on the overall yield of large year classes. This all makes an evaluation for the stock at equilibrium very difficult.

The Working Group is thus unable to provide absolute values for $\mathrm{F}_{\text {msy }}$ or $\mathrm{F}_{\text {msy }}$ proxies, as there are insufficient data to derive absolute estimates of $\mathrm{F}_{\mathrm{msy}}$ with any degree of precision.

There are some additional considerations in relations to exploitation levels to maximise long-term yield, which might indicate that current F might be above $\mathrm{F}_{\text {msy }}$ :

- The stock has a high growth rate with considerable growth potential. Estimates of 0-gp and 1-gp discards are high, thus any improvement in the selectivity pattern would result in increased future yield.
- The age structure is narrow and is not recovering despite a significant decrease in overall effort from the mid-water pelagic fleet.


### 6.3.6 Biological reference points

## Precautionary approach reference points

There is currently no biological basis for defining appropriate reference points, in view of the rapid expansion of the stock size over a short period (ACFM, October 2002). ACFM (2007) proposed that $\mathrm{F}_{\mathrm{pa}}$ be set at 0.5 by association with other haddock stocks, however, the Working Group no longer considers an $\mathrm{F}_{\mathrm{pa}}$ value determined in association with other haddock stocks as appropriate. The absolute level of F in this stock at present is poorly known.

## Yield and biomass-per-recruit

Yield-per-recruit (YPR) and SSB per recruit (SPR) for the Irish Sea stock were calculated by the 2004 WGNSDS, conditional on the exploitation pattern for landings in 2000-2002 given for ages 0 to 5+ by XSA, using MFYPR software. Long-term (19932003) catch weights and stock weights-at-age were used. Input data are given in Ta-
ble 6.3.11, and the summary output is given in Table 6.3.12. The YPR and SPR curves are plotted in Figure 6.3.13. The deterministic output from this model is, however, highly uncertain. Figure 6.3.12 illustrates the uncertainty in the yield-per-recruit curve. Any estimate from the analysis is highly uncertain (high cv values in Table 6.3.10) implying poorly defined $F$ reference point as well as the absolute level of yield.

### 6.3.7 Management plans

There is no specific management plan for haddock in the Irish Sea. Due to the bycatch of cod in the haddock fishery, the regulations affecting Irish Sea haddock remain linked to those implemented under the cod management plan (Council Regulation (EC) $1342 / 2008$ ).

### 6.3.8 Uncertainties and bias in assessment and forecast

This assessment is based on survey trends only as recent levels of catch are uncertain. After a period of poor sampling of landings for length and age, the sampling levels and coverage since 2007 are adequate to allow compilation of catch-at-age data. Discard sampling levels also increased significantly in the last three years. The highly variable and very large estimates of discarding for this fleet observed by previous WG are still evident. Historical landings data for this stock are uncertain, but samplebased estimates of landings suggest that the accuracy of officially reported landings has improved substantially since 2006. The recent catch-at-age data (2003-2006) are still considered too inaccurate, due to poor sampling information, to form the basis for a traditional analytical assessment based on catch-at-age data.

The narrow age range in the haddock stock and the resulting low numbers caught at older ages in the surveys restricted the number of age classes that could be used in the model. This and the differences in catchability-at-age between surveys make the total mortality difficult to estimate. The survey data used in the assessment are quite consistent both internally and between fleets, probably due to the very large data contrast between year-class strengths as well as the restricted distribution of the stock. The recruitment pattern for this stock since the early 1990s is relatively well established and can be tracked fairly consistently through both the surveys and commercial catches. Hence it can be established with some confidence how, qualitatively, the catch and stock is likely to be impacted in the short term by recent year classes.

Knowledge of basic biology of Irish Sea haddock is expanding through data on growth, maturity and distribution obtained during trawl surveys. Patterns of movement within the Irish Sea and between the Irish Sea and surrounding areas are poorly understood, and it is assumed that the Irish Sea stock is essentially self-sustaining at present. Trends in length and weight-at-age in the stock over time are apparent and reduced growth appears to have coincided with the growth of the stock. This may represent density-dependent growth effects (although other environmental factors may contribute) that will affect any forecast and lead to overoptimistic forecast estimates unless correctly predicted.

The projected survey estimate of biomass should only be used for interpreting trends rather than a relative estimate. $\mathrm{F} / \mathrm{Z}$ is poorly estimated and currently unknown. The problem is with using Z-M as a proxy for F in the SURBA-based assessment, when total mortality from the model is poorly defined. The SURBA Z-values are only a relative measure and do not mean anything unless the catchability-at-age in the survey(s) are quantified. The SURBA Z-values cannot be taken as an absolute, which
makes effort based management very difficult, especially measured against a nonstock specific reference point. The additional recruitment survey indices indicate similar above average recruitment in the last year, giving confidence in the higher recruitment indicated by the current survey based assessment. The NIGFS-Oct survey has good internal consistency (see Stock Annex) and both 0-gp indices appear to indicate relative year-class strength well historically (Figure 6.3.2 and 6.3.3).

The perception of the stock from this year's assessment does not differ qualitatively from that obtained last year.

### 6.3.9 Recommendations for next benchmark assessment

The primary concern with this stock is that recent catch-at-age data are considered inaccurate to form the basis for a traditional analytical assessment based on catch-atage data. This has been attributed to poor sampling information, which has improved in the last two years. The absence of reliable discard estimates is also serious deficiency that must be addressed if management is to be based on catch-at-age analysis. Levels of discard sampling have increased substantially in the last three years and reliable discards-at-age matrix could be formulated over the next few years.

The problems in terms of generating reliable catch-at-age numbers for this stock are not likely to be solved in the short term. Furthermore, with the sharp decline in whitefish directed effort in the Irish Sea, sampling opportunities for haddock from landings, are not likely to improve.

### 6.3.10 Management considerations

Following decades of very low recruitment and biomass as indicated by very low fishery catches, this stock grew substantially in the 1990s following sudden pulses of recruitment, and has gone from a minor bycatch species to one of the most economically valuable target species in the Irish Sea. Since the mid-1990s the haddock population in the Irish Sea is experiencing one of the largest and most sustained period of growth. The recruitment signals are clearly revealed by surveys, but the steep age profile in the catches and the resultant dependence of the fishery on highly variable recent year classes means that catch and SSB forecasts will be uncertain. The prevention of directed fishing for haddock during the cod closures in 2000-2010, other than during limited fishing experiments, should have curtailed the directed fisheries on mature haddock that occur in spring.

EU has adopted a long-term plan for cod stocks and the fisheries exploiting those stocks (Council Regulation (EC) 1342/2008). The long-term management plan for cod implemented in the Irish Sea from 2008 will affect catches of species caught in related fisheries, including haddock. The current directed fishery for haddock in the Irish Sea is likely to generate bycatches of cod in the same area.

Sampling schemes since the 1990s have shown high rates of discarding of haddock less than 3 years old and variable discarding of 3-year-olds in fisheries using 7089 mm mesh nets. Samples from whitefish vessels since the introduction of $100+\mathrm{mm}$ mesh and other recent technical measures are too few to form a basis for evaluation of discards in that fleet. Discard rates could be reduced by using more selective fishing gears in the small mesh fisheries. The decline in growth rate might also result in discarding occurring at progressively older ages. However, any measures to reduce discards will result in increased future yield.

Current TAC management measures are not responsive enough considering the dynamic nature of changes in stock abundance. Under the assumption of constant ef-
fort, the increase in abundance from 2005-2008, created increased catch opportunities. During this period the TAC remained relatively constant and resulted in increased discarding of older fish (particularly in 2007). The TAC for 2009 was increased based on the increasing trend of stock abundance, in spite of evidence of weaker recruitment and possible decreasing abundance.

Landings data have not been used in the assessment. Landings data for this stock are uncertain because of species misreporting, which has been estimated from quayside observations in one country only. Restrictive quotas for some countries caused extensive misreporting during the 1990s prior to the introduction of a separate TAC allocation for the Irish Sea. Estimates of misreporting have been included in the estimates of landings, except for 2003. The recent implementation of buyers and sellers legislation has improved the quality of the landings data since 2006.

Under the EU policy for setting TACs, the Irish Sea haddock stock would be classified as a category 9 stock (i.e. state of stock is unknown, but trends based assessment indicates decrease in SSB). The guidelines require firstly an evaluation of current levels of F in relation to FMSY, if a MSY proxy is available. Current F estimates are considered uncertain and unreliable. Survey biomass estimates in the last two years are $>20 \%$ lower than the survey biomass estimates in the previous three years. This category would result in a decrease TAC of $15 \%$.

Table 6.3.1. Nominal landings ( t ) of haddock in Division VIIa, 1984-2009, as officially reported to ICES. (Working Group figures are given in Table 6.3.2).

| Country | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 3 | 4 | 5 | 10 | 12 | 4 | 4 | 1 | 8 | 18 |
| France | 38 | 31 | 39 | 50 | 47 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 73 | 41 |
| Ireland | 199 | 341 | 275 | 797 | 363 | 215 | 80 | 254 | 251 | 252 |
| Netherlands | - | - | - | - | - | - | - | - | - | - |
| UK (England \& Wales) ${ }^{1}$ | 29 | 28 | 22 | 41 | 74 | 252 | 177 | 204 | 244 | 260 |
| UK (Isle of Man) | 2 | 5 | 4 | 3 | 3 | 3 | 5 | 14 | 13 | 19 |
| UK (N. Ireland) | 38 | 215 | 358 | 230 | 196 | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ |
| UK (Scotland) | 78 | 104 | 23 | 156 | 52 | 86 | 316 | 143 | 114 | 140 |
| Total | 387 | 728 | 726 | 1,287 | 747 | 560 | 582 | 616 | 703 | 730 |
|  |  |  |  |  |  |  |  |  |  |  |
| Country | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| Belgium | 22 | 32 | 34 | 55 | 104 | 53 | 22 | 68 | 44 | 20 |
| France | 22 | 58 | 105 | 74 | 86 | $n / a$ | 49 | 184 | 72 | 146 |
| Ireland | 246 | 320 | 798 | 1,005 | 1,699 | 759 | 1,238 | 652 | 401 | 229 |
| Netherlands | - | - | 1 | 14 | 10 | 5 | 2 | - | - | - |
| UK (England \& Wales) ${ }^{1}$ | 301 | 294 | 463 | 717 | 1,023 | 1,479 | 1,061 | 1,238 | 551 | 248 |
| UK (Isle of Man) | 24 | 27 | 38 | 9 | 13 | 7 | 19 | 1 | - | - |
| UK (N. Ireland) | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ |
| UK (Scotland) | 66 | 110 | 14 | 51 | 80 | 67 | 56 | 86 | 47 | 31 |
| Total | 681 | 841 | 1,453 | 1,925 | 3,015 | 2,370 | 2,447 | 2,229 | 1,115 | 674 |


| Country | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 15 | 22 | 23 | 30 | 15 | $7^{*}$ |
| France | 20 | 36 | 20 | 11 | 6 | $\mathbf{-}^{*}$ |
| Ireland | 296 | 139 | 184 | 477 | 319 | $317^{*}$ |
| Netherlands | - | - |  | - | - | - |
| UK (England \& Wales) |  | 421 | 344 | 419 | 559 | 521 |
| UK (Isle of Man) | - | - | - | - | 1 |  |
| UK (N. Ireland) | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ |
| UK (Scotland) | 9 | 6 | 9 | 1 | 17 |  |
| United Kingdom |  |  |  |  |  | $458^{*}$ |
| Total | 761 | 547 | 655 | 1078 | 879 | $782^{*}$ |

*Preliminary.
1989-2008 Northern Ireland included with England and Wales.
n/a = not available .

Table 6.3.2. Haddock in VIIa. Total international landings of haddock from the Irish Sea, 19722009, as officially reported to ICES. Working Group figures, assuming 1972-1992 official landings to be correct, are also given. The 1993-2005 WG estimates include sampled-based estimates of landings at a number of Irish Sea ports. Landings in tonnes live weight.

| Year | Official landings | WG landings |
| :---: | :---: | :---: |
| 1972 | 2204 | 2204 |
| 1973 | 2169 | 2169 |
| 1974 | 683 | 683 |
| 1975 | 276 | 276 |
| 1976 | 345 | 345 |
| 1977 | 188 | 188 |
| 1978 | 131 | 131 |
| 1979 | 146 | 146 |
| 1980 | 418 | 418 |
| 1981 | 445 | 445 |
| 1982 | 303 | 303 |
| 1983 | 299 | 299 |
| 1984 | 387 | 387 |
| 1985 | 728 | 728 |
| 1986 | 726 | 726 |
| 1987 | 1287 | 1287 |
| 1988 | 747 | 747 |
| 1989 | 560 | 560 |
| 1990 | 582 | 582 |
| 1991 | 616 | 616 |
| 1992 | 703 | 656 |
| 1993 | 730 | 813 |
| 1994 | 681 | 1043 |
| 1995 | 841 | 1753 |
| 1996 | 1453 | 3023 |
| 1997 | 1925 | 3391 |
| 1998 | 3015 | 4902 |
| 1999 | 2370 | 4129 |
| 2000 | 2447 | 1380 |
| 2001 | 2229 | 2498 |
| 2002 | 1115 | 1972 |
| 2003 | 674 | n/a |
| 2004 | 761 | 1278 |
| 2005 | 547 | 699 |
| 2006 | 655 | 647 |
| 2007 | 1078 | 1066 |
| 2008 | 879 | 872 |
| 2009 | n/a | 838 |

## Table 6.3.3. Haddock in VIIa: catch numbers-at-age.



Table 6.3.4. Haddock in VIIa: catch weights-at-age.

| Catch weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\mathrm{n} / \mathrm{a}$ | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.351 | 0.346 | 0.361 | 0.346 | 0.348 | 0.19 | 0.325 | 0.329 | 0.3 | 0.279 | $\mathrm{n} / \mathrm{a}$ | 0.401 | 0.273 | 0.244 | 0.240 | 0.300 | 0.306 |
| 2 | 0.596 | 0.56 | 0.545 | 0.474 | 0.592 | 0.53 | 0.416 | 0.474 | 0.452 | 0.357 | n/a | 0.519 | 0.417 | 0.354 | 0.440 | 0.377 | 0.426 |
| 3 | 1.688 | 1.103 | 0.898 | 0.917 | 1.002 | 1.13 | 0.802 | 0.786 | 0.859 | 0.749 | n/a | 1.007 | 0.697 | 0.505 | 0.638 | 0.534 | 0.507 |
| 4 | 2.52 | 2.73 | 1.983 | 2.034 | 1.349 | 2 | 2.064 | 1.573 | 1.243 | 1.361 | $\mathrm{n} / \mathrm{a}$ | 1.940 | 1.256 | 0.872 | 0.786 | 0.743 | 0.778 |
| +gp | 2.52 | 2.522 | 2.178 | 2.682 | 1.955 | 2.55 | 2.854 | 2.365 | 1.869 | 2.107 | n/a | 2.544 | 2.268 | 1.841 | 1.987 | 1.261 | 1.265 |
| 0 SOPCOFAC | 0.9995 | 1.0008 | 1.0007 | 1.0029 | 0.9465 | 0.9958 | 0.9996 | 0.9675 | 1.0002 | 0.9991 |  |  |  |  |  |  |  |

Table 6.3.5. Haddock in VIIa: Estimates of Irish Sea haddock discards 1995-2009. Data are numbers ('000 fish) discarded by the fleet, estimated from numbers per sampled trip raised to total fishing effort by each fleet, for the range of quarters indicated. Tables (b) and (d) represent estimates from limited observer sampling of N.Ireland vessels also included within the self-sampling estimates for N.Ireland trawlers catching Nephrops (Table (a)). Table (f) is the total for sampled fleets and quarters, excluding missing quarters or fleets. Table (e) is the revised figures supplied to the 2005 WG.

|  | 1996 Q1-4 | 1997 Q1-4 | 1998 Q1-4 | 1999 Q1-4 | 2000 Q1-4 | 2001 Q1-4 | 2002 Q1-4 | 2003 Q1 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 43 trips | 39 trips | 48 trips | 39 trips | 44 trips | 43 trips | 35 trips | 8 trips |  |  |  |  |  |  |
|  | 4485 | 100 | 1552 | 1274 | 110 | 1083 | 851 | 0 | n/a | n/a | n/a | n/a |  |  |
|  | 229 | 1209 | 318 | 342 | 2384 | 140 | 1073 | 62 | n/a | n/a | n/a | n/a |  |  |
|  | 179 | 88 | 210 | 69 | 253 | 199 | 37 | 28 | n/a | n/a | n/a | n/a |  |  |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 | n/a | n/a | n/a | n/a |  |  |
| (b) Observer scheme: N.Ireland vessels catching Nephrops (single trawl only) (*not raised to fleet level - no. of fish) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | 1999 Q3-4 | 2000 Q1-3 | 2001 Q1 |  |  |  |  | 2006 Q3-4* | 2007 Q1-4 | 2008 Q1-4 | 2009 Q1-4 |
| Age |  |  |  | 4 trips | 6 trips | 1 trip |  |  |  |  | 9 trips | 29 trips | 55 trips | 30 trips |
|  |  |  |  | 2185 | 210 | 0 |  |  |  |  | 8391 | 901 | 625 | 1609 |
|  |  |  |  | 22 | 280 | 1677 |  |  |  |  | 809 | 1553 | 295 | 284 |
|  |  |  |  | 0 | 57 | 1593 |  |  |  |  | 60 | 681 | 124 | 101 |
|  |  |  |  | 0 | 0 | 0 |  |  |  |  | 15 | 74 | 16 | 23 |
|  |  |  |  | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 1 | 0 |
| (c) Observer scheme: N.Ireland midwater trawl |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 1997 Q2-4 | 1998 Q1-3 | 1999 Q3-4 | 2000 Q1 | 2001 Q1 |  |  |  |  |  |  | 2008 Q4 | 2009 Q2 |
| Age |  | n/a | n/a | 5 trips | 4 trips | 2 trips |  |  |  |  |  |  | 1 trip | 1 trip |
|  |  | 0 | 0 | 68 | 0 | 0 |  |  |  |  |  |  | 0 | 0 |
|  |  | 178 | 316 | 96 | 20 | 0.4 |  |  |  |  |  |  | 7 | 1 |
|  |  | 19 | 1342 | 35 | 83 | 19 |  |  |  |  |  |  | 15 | 39 |
|  |  | 4 | 0 | 2 | 5 | 0 |  |  |  |  |  |  | 2 | 19 |
| (d) Observer scheme: N.Ireland twin trawl (*not raised to fleet level - no. of fish) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 1997 Q2-4 | 1998 Q1-3 | 1999 Q4 | 2000 Q1-4 | 2001 Q1 |  |  |  |  | 2006 Q3-4* | 2007 Q1-4 | 2008 Q1-4 | 2009 Q1-4 |
| Age |  | n/a | n/a | 1 trips | 10 trips | 2 trips |  |  |  |  | 2 trip | 14 trips | 16 trips | 18 trips |
|  |  | 34 | 4 | 26 | 10 | 0 |  |  |  |  | 363 | 369 | 676 | 3219 |
|  |  | 284 | 205 | 3 | 13 | 3 |  |  |  |  | 59 | 275 | 183 | 315 |
|  |  | 6 | 382 | 0 | 10 | 19 |  |  |  |  | 9 | 77 | 70 | 600 |
|  |  | 0.5 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 9 | 6 | 200 |
|  |  | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0 | 0 | 0 | 1 |

Table 6.3.5. (Continued)

|  | 1996 Q1-4 | 1997 Q1-4 | 1998 Q1-4 | 1999 Q1-4 | 2000 Q1-4 | 2001 Q1-4 | 2002 Q1-4 | 2003 Q1-4 | 2004 Q1-4 | 2005 Q1-4 | 2006 Q1-4 | 2007 Q1-4 | 2008 Q1-4 | 2009 Q1-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 8 trips | 8 trips | 7 trips | 4 trips | 10 trips | 2 trips | 1 trip | 9 trips | 11 trips | 8 trips | 5 trips | 16 trips | 18 trips | 18 trips |
| 0 | 3808 | 165 | 565 | 87 | 182 | 5349 | 47 | 1169 | 5663 | 776 | 3966 | 1122 | 322 | 5759 |
| 1 | 713 | 11396 | 1973 | 58 | 2193 | 7354 | 31 | 1747 | 6566 | 2350 | 10140 | 8735 | 1226 | 5654 |
| 2 | 297 | 303 | 3564 | 59 | 580 | 140 | 0 | 1178 | 2301 | 996 | 3856 | 3995 | 783 | 334 |
| 3 | 0 | 0 | 0 | 0 | 0 | 15 | 0 | 10 | 225 | 120 | 132 | 435 | 44 | 72 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 |
| (f) Total for sampled fleets and quarters: NI self sampling scheme (a); NI midwater trawl (c); ROI otter trawl (e) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| Age | 51 trips | n/a | n/a | 48 trips | 58 trips | 47 trips | 36 trips | 17 trips | n/a | n/a | n/a | n/a | n/a | n/a |
| 0 | 8293 | 265 | 2117 | 1429 | 292 | 47 | 36 | 17 | n/a | n/a | n/a | n/a | n/a | n/a |
| 1 | 942 | 12783 | 2607 | 496 | 4597 | 6432 | 898 | 1169 | n/a | n/a | n/a | n/a | n/a | n/a |
| 2 | 476 | 410 | 5116 | 163 | 916 | 7494 | 1104 | 1809 | n/a | n/a | n/a | n/a | n/a | n/a |
| 3 | 0 | 4 | 0 | 2 | 5 | 358 | 37 | 1206 | n/a | n/a | n/a | n/a | n/a | n/a |
| 4 | 0 | 0 | 0 | 0 | 0 | 15 | 11 | 10 | n/a | n/a | n/a | n/a | n/a | n/a |

[^9]Table 6.3.6. Haddock in VIIa: Proportion by number-at-age discarded by sampled fleets.

| Proportion discarded |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Period | age 0 | age 1 | age 2 | age 3 |
| Mid-water trawl | Q2-Q4 1997 |  | 0.93 | 0.37 | 0.02 |
| Mid-water trawl | Q1-Q3 1998 |  | 0.99 | 0.16 | 0.00 |
| Mid-water trawl | Q3-Q4 1999 | 1.00 | 0.79 | 0.31 | 0.00 |
| Mid-water trawl | Q1 2000 |  | 1.00 | 0.44 | 0.04 |
| Mid-water trawl | Q1 2001 |  | 1.00 | 0.30 |  |
| Mid-water trawl | Q4 2008 | 1.00 | 0.97 | 0.90 | 0.30 |
| Mid-water trawl | Q2 2009 |  | - | 0.44 | 0.14 |
| Single Nephrops | Q3-Q4 1999 | 1.00 | 0.94 |  |  |
| Single Nephrops | Q1-Q3 2000 | 1.00 | 0.97 | 0.45 |  |
| Single Nephrops | Q1 2001 |  | 1.00 | 0.49 |  |
| Single Nephrops | Q3-Q4 2006 | 1.00 | 1.00 | 0.96 | 0.50 |
| Single Nephrops | Q1-Q4 2007 | 1.00 | 1.00 | 0.94 | 0.79 |
| Single Nephrops | Q1-Q4 2008 | 1.00 | 0.99 | 0.78 | 0.18 |
| Single Nephrops | Q1-Q4 2009 | 1.00 | 1.00 | 0.88 | 0.46 |
| Twin trawl | Q2-Q4 1997 | 1.00 | 1.00 | 0.61 | 0.04 |
| Twin trawl | Q1-Q3 1998 | 1.00 | 1.00 | 0.76 | 0.00 |
| Twin trawl | Q4 1999 | 1.00 | 1.00 |  |  |
| Twin trawl | Q1 - Q4 2000 | 1.00 | 0.96 | 0.28 |  |
| Twin trawl | Q1 2001 |  | 1.00 | 0.12 |  |
| Twin trawl | Q3-Q4 2006 | 1.00 | 1.00 | 0.81 | 0.00 |
| Twin trawl | Q1-Q4 2007 | 1.00 | 1.00 | 0.91 | 0.63 |
| Twin trawl | Q1-Q4 2008 | 1.00 | 0.95 | 0.50 | 0.05 |
| Twin trawl | Q1-Q4 2009 | 1.00 | 0.99 | 0.95 | 0.75 |
| OTB | Q1-Q4 2007 | 1.00 | 1.00 | 0.93 | 0.65 |
| ОТВ | Q1-Q4 2008 | 1.00 | 0.97 | 0.90 | 0.17 |
| ОТВ | Q1-Q4 2009 | 1.00 | 1.00 | 0.62 | 0.24 |

Table 6.3.7. Haddock in VIIa: stock weights-at-age.

| Stock weights-at-age (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0.095 | 0.083 | 0.085 | 0.083 | 0.070 | 0.060 | 0.057 | 0.048 | 0.051 | 0.056 | 0.050 | 0.041 | 0.032 | 0.035 | 0.034 | 0.041 | 0.044 | 0.039 |
| 2 | 0.425 | 0.342 | 0.352 | 0.365 | 0.361 | 0.257 | 0.228 | 0.232 | 0.204 | 0.218 | 0.233 | 0.201 | 0.167 | 0.130 | 0.144 | 0.142 | 0.170 | 0.186 |
| 3 | 1.073 | 0.977 | 0.792 | 0.796 | 0.873 | 0.749 | 0.567 | 0.514 | 0.551 | 0.476 | 0.489 | 0.515 | 0.462 | 0.383 | 0.302 | 0.326 | 0.332 | 0.395 |
| 4 | 1.794 | 2.043 | 1.709 | 1.318 | 1.436 | 1.388 | 1.292 | 0.967 | 0.929 | 0.979 | 0.798 | 0.816 | 0.904 | 0.801 | 0.682 | 0.517 | 0.589 | 0.586 |
| +gp | 2.589 | 3.062 | 3.149 | 2.513 | 2.171 | 2.033 | 2.149 | 1.976 | 1.633 | 1.494 | 1.429 | 1.202 | 1.269 | 1.373 | 1.298 | 1.060 | 0.875 | 0.916 |

Table 6.3.8. Haddock in VIIa: Available tuning data (file name: h7ani.tun).
IRISH SEA haddock, 2010 WG, ANON, COMBSEX, TUNING DATA(effort, nos at age)
104
NIGFS March
19922010
110.210 .25

15

| 1 | 1525 | 23 | 0 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 139 | 569 | 31 | 0 | 0 | 0 |
| 1 | 644 | 58 | 183 | 0 | 0 | 0 |
| 1 | 24823 | 437 | 0 | 43 | 0 | 0 |
| 1 | 1065 | 3743 | 67 | 3 | 1 | 0 |
| 1 | 25118 | 474 | 1457 | 44 | 0 | 2 |
| 1 | 3913 | 8694 | 70 | 105 | 1 | 0 |
| 1 | 6058 | 680 | 2072 | 16 | 11 | 0 |
| 1 | 14028 | 1853 | 64 | 147 | 2 | 3 |
| 1 | 3277 | 6990 | 770 | 40 | 20 | 0 |
| 1 | 28755 | 842 | 1059 | 78 | 1 | 0 |
| 1 | 6966 | 14162 | 341 | 356 | 26 | 0 |
| 1 | 19945 | 2379 | 2206 | 45 | 35 | 0 |
| 1 | 24488 | 6454 | 406 | 234 | 13 | 2 |
| 1 | 13444 | 12721 | 2194 | 91 | 33 | 0 |
| 1 | 20918 | 11325 | 3661 | 240 | 16 | 11 |
| 1 | 7480 | 12009 | 2559 | 495 | 48 | 0 |
| 1 | 9345 | 3888 | 2877 | 163 | 37 | 5 |
| 1 | 17058 | 1765 | 524 | 239 | 26 | 1 |

Fleets below not included in assessment
NIGFS Oct
19912009
110.830 .88

03

| 1 | 15780 | 70 | 0 | 0 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 124 | 784 | 151 | 0 | 0 | 0 | 0 |
| 1 | 4462 | 101 | 375 | 3 | 0 | 0 | 0 |
| 1 | 56683 | 1137 | 12 | 79 | 0 | 0 | 1 |
| 1 | 1661 | 10153 | 74 | 0 | 5 | 0 | 0 |
| 1 | 143300 | 1167 | 1480 | 13 | 0 | 0 | 0 |
| 1 | 16400 | 39680 | 174 | 98 | 1 | 0 | 0 |
| 1 | 41820 | 1243 | 3778 | 22 | 3 | 4 | 0 |
| 1 | 80674 | 2835 | 71 | 145 | 0 | 1 | 0 |
| 1 | 6545 | 8598 | 763 | 31 | 39 | 0 | 0 |
| 1 | 75017 | 2003 | 2742 | 311 | 0 | 20 | 0 |
| 1 | 15116 | 10501 | 86 | 365 | 0 | 0 | 0 |
| 1 | 53922 | 7125 | 3008 | 59 | 79 | 0 | 0 |
| 1 | 70337 | 14413 | 1261 | 649 | 0 | 0 | 0 |
| 1 | 47030 | 12962 | 1743 | 59 | 8 | 0 | 0 |
| 1 | 35748 | 10788 | 3607 | 392 | 52 | 0 | 0 |
| 1 | 9654 | 9804 | 4050 | 1057 | 41 | 0 | 0 |
| 1 | 9037 | 4880 | 2242 | 277 | 24 | 0 | 0 |
| 1 | 45869 | 4269 | 951 | 459 | 29 | 12 | 3 |

MIK net May/June
19942009
110.380 .47

00

| 1 | 47000 |
| ---: | ---: |
| 1 | 1700 |
| 1 | 47800 |
| 1 | 14500 |
| 1 | 2500 |
| 1 | 15400 |
| 1 | 1700 |
| 1 | 17100 |
| 1 | 1200 |
| 1 | 4250 |
| 1 | 25970 |
| 1 | 8250 |
| 1 | 40240 |
| 1 | 3820 |
| 1 | 6638 |
| 1 | 18540 |

Table 6.3.9. Haddock in VIIa: SURBA 3.0 fitted numbers-at-age, total mortality-at-age, SSB and Z using the NIGFS-Mar survey data.

| Numbers-at-age |  |  |  |  |  | Total mortality-at-age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  |  |  |  | Age |  |  |  |  |
| Year | 1 | 2 | 3 | 4 | 5 | 1 | 2 | 3 | 4 | 5 |
| 1992 | 0.357 | 0.013 | 0 | 0 | 0 | 0.667 | 0.662 | 1.110 | 1.359 | 1.359 |
| 1993 | 0.056 | 0.183 | 0.007 | 0 | 0 | 0.850 | 0.844 | 1.416 | 1.733 | 1.733 |
| 1994 | 0.399 | 0.024 | 0.079 | 0.002 | 0 | 1.019 | 1.011 | 1.697 | 2.077 | 2.077 |
| 1995 | 5.815 | 0.144 | 0.009 | 0.014 | 0 | 1.346 | 1.335 | 2.240 | 2.742 | 2.742 |
| 1996 | 0.455 | 1.514 | 0.038 | 0.001 | 0.001 | 0.917 | 0.909 | 1.526 | 1.868 | 1.868 |
| 1997 | 9.260 | 0.182 | 0.610 | 0.008 | 0 | 1.270 | 1.260 | 2.114 | 2.588 | 2.588 |
| 1998 | 0.740 | 2.601 | 0.052 | 0.074 | 0.001 | 1.253 | 1.243 | 2.086 | 2.553 | 2.553 |
| 1999 | 2.919 | 0.211 | 0.751 | 0.006 | 0.006 | 1.208 | 1.199 | 2.012 | 2.463 | 2.463 |
| 2000 | 5.612 | 0.872 | 0.064 | 0.100 | 0.001 | 1.110 | 1.101 | 1.849 | 2.263 | 2.263 |
| 2001 | 1.234 | 1.849 | 0.290 | 0.010 | 0.010 | 1.238 | 1.228 | 2.061 | 2.523 | 2.523 |
| 2002 | 6.885 | 0.358 | 0.541 | 0.037 | 0.001 | 0.819 | 0.813 | 1.364 | 1.670 | 1.670 |
| 2003 | 2.141 | 3.035 | 0.159 | 0.138 | 0.007 | 1.008 | 1.000 | 1.678 | 2.054 | 2.054 |
| 2004 | 6.912 | 0.781 | 1.117 | 0.030 | 0.018 | 1.128 | 1.119 | 1.878 | 2.299 | 2.299 |
| 2005 | 10.369 | 2.237 | 0.255 | 0.171 | 0.003 | 1.108 | 1.099 | 1.844 | 2.257 | 2.257 |
| 2006 | 6.795 | 3.425 | 0.746 | 0.040 | 0.018 | 0.946 | 0.939 | 1.575 | 1.928 | 1.928 |
| 2007 | 9.745 | 2.638 | 1.340 | 0.154 | 0.006 | 0.986 | 0.978 | 1.642 | 2.010 | 2.010 |
| 2008 | 3.354 | 3.635 | 0.992 | 0.259 | 0.021 | 1.223 | 1.214 | 2.037 | 2.493 | 2.493 |
| 2009 | 2.847 | 0.987 | 1.080 | 0.129 | 0.021 | 1.285 | 1.275 | 2.140 | 2.619 | 2.619 |
| 2010 | 6.209 | 0.788 | 0.276 | 0.127 | 0.009 | 1.165 | 1.156 | 1.940 | 2.374 | 2.374 |

Stock summary

| Year | Recruits <br> (age 1) | log SE <br> (rec) | SSB | TSB | Z(2-3) | SE (Z) |
| :--- | :---: | :--- | :--- | :--- | :--- | :--- |
| 1992 | 0.357 | 0.360 | 0.006 | 0.040 | 0.886 | 0.379 |
| 1993 | 0.056 | 0.294 | 0.085 | 0.090 | 1.130 | 0.273 |
| 1994 | 0.399 | 0.265 | 0.088 | 0.122 | 1.354 | 0.213 |
| 1995 | 5.815 | 0.285 | 0.083 | 0.577 | 1.788 | 0.185 |
| 1996 | 0.455 | 0.246 | 0.586 | 0.624 | 1.218 | 0.209 |
| 1997 | 9.260 | 0.261 | 0.610 | 1.258 | 1.687 | 0.178 |
| 1998 | 0.740 | 0.259 | 0.811 | 0.855 | 1.664 | 0.174 |
| 1999 | 2.918 | 0.257 | 0.494 | 0.661 | 1.606 | 0.173 |
| 2000 | 5.612 | 0.250 | 0.333 | 0.602 | 1.475 | 0.176 |
| 2001 | 1.234 | 0.266 | 0.563 | 0.626 | 1.645 | 0.175 |
| 2002 | 6.885 | 0.237 | 0.373 | 0.759 | 1.088 | 0.180 |
| 2003 | 2.140 | 0.247 | 0.905 | 1.012 | 1.339 | 0.180 |
| 2004 | 6.912 | 0.253 | 0.778 | 1.061 | 1.499 | 0.176 |
| 2005 | 10.369 | 0.253 | 0.650 | 0.981 | 1.471 | 0.174 |
| 2006 | 6.795 | 0.244 | 0.788 | 1.026 | 1.257 | 0.178 |
| 2007 | 9.745 | 0.253 | 0.897 | 1.229 | 1.310 | 0.180 |
| 2008 | 3.354 | 0.281 | 0.995 | 1.133 | 1.625 | 0.177 |
| 2009 | 2.847 | 0.321 | 0.621 | 0.746 | 1.707 | 0.180 |
| 2010 | 6.209 | 0.397 | 0.339 | 0.581 | 1.548 | 0.110 |

Table 6.3.10. Haddock VIIa: Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships and the yield-per-recruit $F_{\text {msy }}$ proxies.
Stock name
Had-7a
Sen filename
had-7a.sen
pf, pm
Number of iterations $\quad 0$
$\quad 1000$
Simulate variation in Biological parameters
$\quad$ TRUE
SR relationship constrained
TRUE

## Ricker

767/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 1.45 | 0.46 | 4629 | 2523 | 1.15 | 0.30 | 4.04 | 0.00022 | 34.25 |
| Mean | 1.36 | 0.55 | 7784 | 4833 | 1.70 | 0.44 | 8.15 | 0.00033 |  |
| 5\%ile | 0.44 | 0.21 | 1594 | 1414 | 0.74 | 0.07 | 2.29 | $5.00 \mathrm{E}-05$ |  |
| 25\%ile | 0.72 | 0.33 | 2507 | 2195 | 1.07 | 0.24 | 3.65 | 0.00018 |  |
| 50\%ile | 1.07 | 0.47 | 3441 | 2778 | 1.42 | 0.42 | 5.49 | 0.00031 |  |
| 75\%ile | 1.68 | 0.65 | 5575 | 3732 | 2.02 | 0.60 | 8.96 | 0.00044 |  |
| 95\%ile | 3.36 | 1.22 | 17254 | 8047 | 3.43 | 0.93 | 21.81 | 0.0007 |  |
| CV | 0.67 | 0.62 | 4.86 | 5.25 | 0.61 | 0.61 | 1.13 | 0.61 |  |

Beverton-Holt
813/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 2.80 | 0.29 | 7030 | 2580 | 0.44 | 0.80 | 7964 | 1111 | 34.12 |
| Mean | 1.15 | 0.20 | 58936 | 9346 | 0.45 | 1.31 | 41130 | 22121 |  |
| 5\%ile | 0.31 | 0.07 | 2363 | 848 | 0.05 | 0.63 | 3484 | 153 |  |
| 25\%ile | 0.51 | 0.14 | 4913 | 1657 | 0.22 | 0.89 | 5903 | 1014 |  |
| 50\%ile | 0.82 | 0.19 | 9186 | 2574 | 0.38 | 1.12 | 9186 | 2705 |  |
| 75\%ile | 1.46 | 0.25 | 19246 | 4389 | 0.59 | 1.45 | 16093 | 6579 |  |
| 95\%ile | 3.15 | 0.36 | 129006 | 17393 | 1.00 | 2.31 | 70557 | 40158 |  |
| CV | 0.82 | 0.43 | 7.6 | 8.4 | 1.27 | 0.80 | 11.25 | 13.45 |  |

Smooth hockeystick
918/1000 Iterations resulted in feasible parameter estimates

|  | Fcr |  | Fmsy |  | Bmsy |  | MSY |  | ADMB | Alpha | ADMB | Beta | Uns | aled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic |  | 0.87 |  | 0.41 |  | 5359 |  | 2661 |  | 0.49 |  | 0.92 |  | 1.27 | 2727 | 34.55 |
| Mean |  | 0.90 |  | 0.38 |  | 10384 |  | 3359 |  | 0.60 |  | 0.99 |  | 1.56 | 2941 |  |
| 5\%ile |  | 0.33 |  | 0.14 |  | 2439 |  | 1534 |  | 0.30 |  | 0.49 |  | 0.78 | 1439 |  |
| 25\%ile |  | 0.50 |  | 0.28 |  | 3943 |  | 2304 |  | 0.43 |  | 0.66 |  | 1.13 | 1960 |  |
| 50\%ile |  | 0.69 |  | 0.37 |  | 5546 |  | 3010 |  | 0.56 |  | 0.95 |  | 1.45 | 2797 |  |
| 75\%ile |  | 1.04 |  | 0.47 |  | 8645 |  | 4073 |  | 0.71 |  | 1.30 |  | 1.85 | 3830 |  |
| 95\%ile |  | 2.05 |  | 0.66 |  | 22638 |  | 6218 |  | 1.06 |  | 1.64 |  | 2.76 | 4840 |  |
| CV |  | 0.77 |  | 0.42 |  | 2.44 |  | 0.48 |  | 0.41 |  | 0.38 |  | 0.41 | 0.38 |  |
| Per recruit |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | F35 |  | F40 |  | F01 |  | Fmax |  | Bmsypr |  | MSYp |  | Fpa |  | Flim |  |
| Deterministic |  | 0.24 |  | 0.20 |  | 0.20 |  | 0.41 |  | 0.77 |  | 0.38 |  | 0 | 0 |  |
| Mean |  | 0.20 |  | 0.17 |  | 0.18 |  | 0.39 |  | 1.20 |  | 0.39 |  |  |  |  |
| 5\%ile |  | 0.05 |  | 0.04 |  | 0.05 |  | 0.15 |  | 0.39 |  | 0.28 |  |  |  |  |
| 25\%ile |  | 0.15 |  | 0.12 |  | 0.14 |  | 0.29 |  | 0.55 |  | 0.34 |  |  |  |  |
| 50\%ile |  | 0.20 |  | 0.17 |  | 0.19 |  | 0.38 |  | 0.71 |  | 0.38 |  |  |  |  |
| 75\%ile |  | 0.26 |  | 0.22 |  | 0.23 |  | 0.48 |  | 0.97 |  | 0.44 |  |  |  |  |
| 95\%ile |  | 0.34 |  | 0.29 |  | 0.29 |  | 0.67 |  | 2.20 |  | 0.55 |  |  |  |  |
| CV |  | 0.44 |  | 0.43 |  | 0.39 |  | 0.43 |  | 2.06 |  | 0.22 |  |  |  |  |

Table 6.3.11. Haddock in VIIa: Input for yield/Recruit.
MFYPR version 2a
Run: Had7a_2004WG_yield
Had7a_2004WG_yieldMFYPR Index file 11/05/2004
Time and date: 10:55 13/05/2004
Fbar age range: 2-4

| Age | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.2 | 0 | 0 | 0 | 0.000 | 0.000 | 0.000 |
| 1 | 0.2 | 0 | 0 | 0 | 0.061 | 0.140 | 0.322 |
| 2 | 0.2 | 1 | 0 | 0 | 0.302 | 0.544 | 0.492 |
| 3 | 0.2 | 1 | 0 | 0 | 0.754 | 1.118 | 0.967 |
| 4 | 0.2 | 1 | 0 | 0 | 1.377 | 1.057 | 1.814 |
| 5 | 0.2 | 1 | 0 | 0 | 2.259 | 1.057 | 2.308 |

Weights in kilograms

Table 6.3.12. Haddock in VIIa: Yield-per-recruit output table.

MFYPR version 2 a
Run: Had7a_2004WG_yield
Time and date: 10:55 13/05/2004
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 5.8695 | 3.6979 | 5.8200 | 3.6979 | 5.8200 |
| 0.1000 | 0.0906 | 0.2211 | 0.3492 | 4.4167 | 3.5229 | 2.5980 | 3.4733 | 2.5980 |  |
| 0.2000 | 0.1813 | 0.3298 | 0.4658 | 3.8781 | 2.4296 | 2.0593 | 2.3801 | 2.0593 |  |
| 0.3000 | 0.2719 | 0.3951 | 0.5037 | 3.5564 | 1.8139 | 1.7377 | 1.7644 | 1.7377 |  |
| 0.4000 | 0.3626 | 0.4390 | 0.5098 | 3.3412 | 1.4279 | 1.5225 | 1.3783 | 1.5225 | 2.3801 |
| 0.5000 | 0.4532 | 0.4709 | 0.5022 | 3.1861 | 1.1681 | 1.3674 | 1.1186 | 1.3674 | 1.7644 |
| 0.6000 | 0.5439 | 0.4952 | 0.4888 | 3.0683 | 0.9843 | 1.2496 | 0.9347 | 1.2496 | 1.1186 |
| 0.7000 | 0.6345 | 0.5146 | 0.4735 | 2.9752 | 0.8490 | 1.1564 | 0.7995 | 1.1564 | 0.9347 |
| 0.8000 | 0.7252 | 0.5305 | 0.4580 | 2.8993 | 0.7464 | 1.0805 | 0.6969 | 1.0805 | 0.6995 |
| 0.9000 | 0.8158 | 0.5438 | 0.4431 | 2.8358 | 0.6666 | 1.0171 | 0.6170 | 1.0171 | 0.6170 |
| 1.0000 | 0.9065 | 0.5552 | 0.4293 | 2.7818 | 0.6030 | 0.9631 | 0.5535 | 0.9631 | 0.5535 |
| 1.1000 | 0.9971 | 0.5651 | 0.4167 | 2.7350 | 0.5515 | 0.9163 | 0.5019 | 0.9163 | 0.5019 |
| 1.2000 | 1.0878 | 0.5739 | 0.4052 | 2.6939 | 0.5090 | 0.8751 | 0.4594 | 0.8751 | 0.4594 |
| 1.3000 | 1.1784 | 0.5817 | 0.3947 | 2.6573 | 0.4733 | 0.8386 | 0.4238 | 0.8386 | 0.4238 |
| 1.4000 | 1.2691 | 0.5887 | 0.3853 | 2.6245 | 0.4431 | 0.8057 | 0.3936 | 0.8057 | 0.3936 |
| 1.5000 | 1.3597 | 0.5951 | 0.3768 | 2.5947 | 0.4172 | 0.7760 | 0.3676 | 0.7760 | 0.3676 |
| 1.6000 | 1.4503 | 0.6009 | 0.3692 | 2.5676 | 0.3946 | 0.7489 | 0.3451 | 0.7489 | 0.3451 |
| 1.7000 | 1.5410 | 0.6063 | 0.3622 | 2.5427 | 0.3749 | 0.7240 | 0.3253 | 0.7240 | 0.3253 |
| 1.8000 | 1.6316 | 0.6113 | 0.3559 | 2.5197 | 0.3574 | 0.7010 | 0.3079 | 0.7010 | 0.3079 |
| 1.9000 | 1.7223 | 0.6159 | 0.3501 | 2.4983 | 0.3418 | 0.6796 | 0.2923 | 0.6796 | 0.2923 |
| 2.0000 | 1.8129 | 0.6202 | 0.3449 | 2.4784 | 0.3278 | 0.6597 | 0.2783 | 0.6597 | 0.2783 |


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(2-4) | 1.0000 | 0.9065 |
| FMax | 0.3811 | 0.3455 |
| F0.1 | 0.2074 | 0.188 |
| F35\%SPR | 0.2494 | 0.2261 |
|  |  |  |
| Weights in kilograms |  |  |



Figure 6.3.1. Haddock in VIIa: Growth of haddock in the Irish Sea. Top two panels: mean length-at-age in UK(NI) groundfish surveys in March, by year and age, and expected mean weight-atlength based on length-weight parameters from each survey. Lower panels: mean length-at-age from March surveys, and from Quarter 1 commercial landings at age 3 and over, by year class. Lines are Von Bertalanffy model fits with year-class effect included. Model residuals are shown for the fit without year-class effects, and for the fit with year-class effects.


Figure 6.3.2. Haddock in VIIa: Trends in raw survey indices compared with international landings, by age class and year. All values are standardised to the mean for years common to all series in each plot (except for short FSP series).


Figure 6.3.3. Haddock in VIIa: Time-series plots of the logarithms of survey indices at age by year class, after standardising by dividing by the series mean for years from 1991. Data have only been illustrated for the most abundant ages for comparison of year-class signals.


Figure 6.3.4. Haddock in VIIa: Comparison in the relative trends of SSB form 2009 SURBA run and the Irish Sea annual egg production method survey estimates of SSB +2 SE) (Armstrong et al., WD11).


Figure 6.3.5. Haddock in VIIa: Mean Standardised empirical SSB indices from the NIGFS-Mar and NIGFS-Oct surveys, based on raw indices up to age 6 .


Figure 6.3.6. Haddock VIIa: SURBA 3.0 Residuals at age (top panel) and retrospective plots (bottom panel) for the NIGFS-Mar survey.


Figure 6.3.7. Haddock VIIa: Summary plots of landings and results of final SURBA 3.0 run using the NIGFS-Mar survey data. Dotted lines are +/- 1SE. Empirical estimates of SSB and Z given by SURBA from the raw survey data are also shown.


Figure 6.3.8. Haddock VIIa: SURBA 3.0 Residuals-at-age for final run using the NIGFS-Mar survey data.


Figure 6.3.9. Haddock VIIa: Trends in SSB, recruitment and Z(2-3) from the 2009 and 2010 SURBA. SSB and recruitment are standardised to the mean for years common to all series (19922009) in each plot.


Figure 6.3.10. Haddock VIIa: Trend in SSB form 2010 SURBA projected to 2011 (top panel) and SURBA estimate of recruitment compared to available 0-gp indices. SSB and recruitment are standardised to the mean for years common to all series (1994-2009) in each plot.


Figure 6.3.11. Haddock VIIa: MSY fitted stock and recruitment relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $\mathbf{F}_{\text {msy. }}$. Right hand panels: curves plotted from the first 100 MCMC re-samples with converged $\mathbf{F}_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of $\mathrm{Fmsy}_{\text {from }}$ from the $\mathbf{1 0 0 0}$ re-samples.

Had-7a - Per recruit statistics


Figure 6.3.12. Haddock VIIa: Fitted yield-per-recruit F reference points, yield-per-recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric resampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first 100 re-samples.


MFYPR version 2 a
Run: Had7a_2004WG_yield
Time and date: 10:55 13/05/2004

| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(2-4) | 1.0000 | 0.9065 |
| FMax | 0.3811 | 0.3455 |
| F0.1 | 0.2074 | 0.1880 |
| F35\%SPR | 0.2494 | 0.2261 |
|  |  |  |
| Weights in kilograms |  |  |

Figure 6.3.13. Haddock VIIa: Yield-per-recruit based on analysis carried out in 2004.

### 6.4 Nephrops in Division VIIa (Irish Sea East, FU14)

## Type of assessment in 2010

The assessment determines the health of the stock by looking at trends in total landings, lpue, size composition, and biological data from the commercial fisheries. For the first time for this stock the results from UWTV survey data are used to calculate absolute abundance estimates for 2009 and catch options following the process benchmarked at WKNEPH (2009).

## ICES advice applicable to 2009

This stock was reassessed in 2008 based on trends in the fishery and biological parameters. The advice for this fishery for 2009 and 2010 was that landings and effort should not increase above that recorded for 2007.

## ICES advice applicable to 2010

The advice was biannual and still valid from the 2008 assessment which implied that effort should not increase compared to 2007 levels.

### 6.4.1 General

## Stock description and management units

The Irish Sea East Nephrops stock (FU14) is in ICES Subarea VII which includes the Irish Sea West (FU15) stock; the Porcupine Bank (FU16); Aran Grounds (FU17); North-West Irish Coast (FU18), South-East and South-West Irish Coast (FU19); and the Celtic Sea stock (FU20-22). The TAC is set for the whole of Subarea VII which does not correspond to the areas occupied by these stocks.


[^10]
## Management applicable in 2009 and 2010

The TAC is currently set for the larger TAC Area VII. The TAC for 2010 is currently set at 22432 t , a $9 \%$ reduction on the 2009 TAC of 24650 t and $11 \%$ reduction on 2007/2008 TAC of 25 153. The TAC area includes a number of Nephrops stocks showing different levels of exploitation. A single TAC covering a number of distinct stocks allows the possibility of unrestricted catches being taken from a heavily exploited stock when advice suggests they should be limited.

In 2009 the main fleets targeting Nephrops include directed single-rig and twin-rig otter trawlers operating out of ports in UK (NI), UK (E\&W) and Ireland. Details of all regulations including effort controls in place are provided in the Stock Annex.

## The fishery in 2009

Between 1999 and 2003 the number of vessels fishing for Nephrops in FU14 declined by $40 \%$ to a fleet of around 50 vessels. This was largely due to the reduction in the number of visiting UK vessels and the decommissioning of part of the Northern Ireland and local English fleets. Since then, the number of vessels fishing the area has returned to and settled at around 80 vessels over the last three years, mainly from Northern Ireland. Currently, just under 30 of these vessels, between 9 and 21 m in length, have their 'home' ports in Whitehaven, Maryport and Fleetwood, England. The rest of the fleet is generally made up of larger vessels from Northern Ireland.

In 2009 about $70 \%$ of the landings from this fishery were made to Whitehaven and about $20 \%$ to Kilkeel. Over half of the Northern Ireland and a few of the English vessels use twin or triple trawls and account for around $30 \%$ of the Nephrops landings in weight from this FU. Between 1999 and 2009, the recorded number of vessels using these multiple trawls has fluctuated without trend between 15 and 29 vessels, with around $90 \%$ of these vessels coming from Northern Ireland. The earlier decline in the fleet was mainly in the number of single trawlers.

Of the Northern Ireland fleet the proportion returning at the end of a Nephrops trip in FU14, to land in to Northern Ireland, increased from only 6\% in 1999 to around 30\% from 2005 onward.

There has been little apparent change in the make-up of the English and Welsh fleet over the last three of years. However the current state of other stocks, technical conservation and cod recovery measures has had an effect on mesh sizes and fishing patterns. The number of recorded trips has increased yet effort has effectively declined. The average days per trip has declined to a value in 2009 where over $60 \%$ of the trips reported were one day trips. Traditionally a summer fishery, anecdotal data and records of monthly landings indicate the season is starting earlier and ending earlier.

In 2009 the most productive period in this fishery was June to July due in part to poor catches of prawns in the previous period but often due to the weather. The local enforcement agency at Whitehaven recorded that fishing over this period was better than had been experienced for some time with good landings of good sized prawns. The larger Northern Ireland vessels continued to make larger landings and fish predominately further offshore while the local fleet tended to fish closer inshore. Both weather and landings were poor for most of August and poor catches persisted. By September some of the UK fleet had already moved to other fisheries; including the North Sea. The number of UK vessels moving from this summer fishery to the Farn Deeps fishery in winter dropped from 30 in 2007 to only 9 in 2008 but increased again in 2009 to 17.

### 6.4.2 Data available

An overview of the data provided and used by the WG is provided in Table 2.1.

## Landings

Official landings as reported to ICES from FU14 are presented in Table 6.4.1 and were updated for 2009. Between 1987 and 2006 landings from FU14 appeared relatively stable, fluctuating around a long-term average of about 550 t (Table 6.4.2 and Figure 6.4.1). Landings in 2007 were at their highest level since 1978 at 959 t , this is after landings dropped in 2003 to their lowest apparent level since 1974. The landings figure declined to 676 t in 2008 and rose slightly to 694 t in 2009. These landings are still higher than any others recorded since 1991; however this could be due to change in the process for recording landings. The introduction of the buyers and sellers legislation in 2006 by the UK precludes direct comparison with previous years as reported levels are considered to have significantly improved.

Over the last 10 years UK vessels have landed, on average, $87 \%$ of the reported annual international landings. Irish vessels increased their share of the landings to $35 \%$ in 2002 but it has since declined to $2 \%$ in 2009 (Table 6.4.2).

## Length composition

Quarterly length compositions of landings, catch and discards were available from the UK England and Wales for most of the period 1992-2009. The numbers of samples taken are presented in Table 2.1. The raising and collation procedures are documented in the Stock Annex B1. Landings sampling deteriorated in 2005, it improved in 2007 and have remained at a consistently high level since. Figure 6.4.4 shows the annual catch and landings length distributions. Discard rates have been estimated from the same figures and have declined in the last six years from $24 \%$ to $4 \%$ of total catch by weight and $43 \%$ and $8 \%$ by number. Females generally have a higher discard rate because they are generally smaller. The sharp decline in the discard rate from 2008 to 2009 particularly for males might suggest a change in discard practice but the shift to the right for the catch distribution in 2009 and the minimum observed size might suggest a decline in recruitment. This could be partly a sampling artefact as only 10 observer trips were carried out in 2009, around a third of the number carried out in 2008. These observer trips have been the only source for catch and discard data in recent years. The landings were still well sampled so these concerns are only limited to defining the discarded component of the catch in 2009. A summary of mean size information is provided in Table 6.4.5. In 2009 the local enforcement agency remarked on improved catches of good sized prawns and better fishing than had been seen for some time, which would support the observed shift to larger prawns in the catch and a small increase in mean landing size of the males.

## Commercial cpue

A $10 \%$ TAC increase in 2006 followed by a $17 \%$ increase in 2007 coupled with the implementation in the UK of buyers and sellers regulations effective from and throughout 2006, has improved the accuracy of reported landings information. This appears to have reduced the reasons to misreport, despite the decline in TAC 2009 for Area VII, and the legislation provides the quality control. Landings have not exceeded the advised TAC for this Functional Unit.

The introduction of the buyers and sellers legislation for 2006 complicates the interpretation of any prior trends. In 2009, most of the landings were made into England with a high proportion of these landings ( $60 \%$ of the directed landings) being made
by visiting Northern Irish vessels. UK Nephrops directed effort fluctuated around a downward trend since 1978 reaching a minimum in 2004. Since then effort has remained relatively stable fluctuating without trend around a mean of 13400 hrs . The effort for this fleet in 2009 was recorded at its lowest in the series at 12000 hrs . Quarterly effort plots show a predominance of effort in the 2 nd and 3rd quarters (Figure 6.4.2).

In light of the limited indices available for this stock, trends in recent lpue are still reviewed as some of the best available information despite reservations about the accuracy of the historical landings. The UK lpue series is based on a combination of directed Nephrops voyages by English and Welsh vessels landing to Fleetwood and Whitehaven, where the weight of Nephrops landed is more than $25 \%$ of the total landing and all trips by visiting Northern Irish vessels which target Nephrops (Table 6.4.4). Analysis of the lpue trends for this reference fleet shows that between 1989 and 2004 there is little correspondence between the E\&W and NI figures. Uncertainties about the recorded landings during this period could account for some of the differences as they fall back into step after 2004. Further data and analysis is required to determine whether this series continues to be appropriate. Between 1990 and 2003 the combined lpue has fluctuated between 17 and $26 \mathrm{~kg} /$ hour trawling. Since then lpue has risen year on year to $40 \mathrm{~kg} /$ hour trawling in 2007, the highest level in the series (Figure 6.4.1). Since then the annual lpue has effectively stayed at this level. In 2009 the annual lpue was at $39 \mathrm{~kg} /$ hour. The fluctuations over the last three years reflect the influence of the NI fleet on the series as the lpue of the E\&W fleet has continued to rise. The lpue of the Northern Irish fleet is driving this trend and since 2004 has been at a level comparable to the Republic of Ireland fleet. This could reflect a change in reporting and/or a change in targeted effort rather than any biological phenomena.

Male Nephrops predominate landings and the annual proportion of females appears highly dependent on the fishing effort in the third quarter (Figure 6.4.2). Lpues for males and females $<35 \mathrm{~mm}$ CL (Figure 6.4.3) appear to exhibit the same general trends. Minima in 2003 were followed by upward trends to the highest values in both series in 2007. They have both since declined but still remain above any other values in the series. The lpue of the larger males ( $>35 \mathrm{~mm}$ ) has been increasing since 2002. Whether this is an artefact of the recording practice is unclear but since the improvement in 2006 it has continued to rise. The quarterly pattern of availability to the fishery of females $>35 \mathrm{~mm}$, means that meaningful statistics for this portion of the population are highly dependent upon the level of fishing and the sampling effort deployed in the 3rd quarter.

The increasing lpue of the $<35 \mathrm{~mm}$ CL categories up to 2007 and decline in mean size of the landings (Figure 6.4.3 and Figure 6.4.1) and the increase in the range of sizes in the catch (Figure 6.4.4) could be indicative of good recruitment. This was supported by the local enforcement agency who at the time noted an increase in the proportion of tails landed. The trends since have been reversed suggesting a decline in recruitment.

## Surveys

In August 2007, 2008 and 2009 the UK and the Republic of Ireland carried out a joint underwater TV survey of the Nephrops grounds in the Eastern Irish Sea. The survey was of a fixed grid design and was carried out using the same protocols used in UWTV surveys in the Western Irish Sea. This survey and stock was not reviewed at WKNEPH 2009 but the protocols and standardised process has been adopted see Stock Annex.

In 2007 poor visibility hampered the survey and despite repeated attempts at over 15 stations, turbidity scores precluded the use of some of the counts. On first analysis only 20 stations were initially considered usable, a recent review of this data suggest that the original analysis was over optimistic. The 2008 and 2009 surveys were both far more successful. A new camera and sledge improved the resolution of the footage captured. Sea conditions were far better so the quality of the video data collected was much improved ( 35 and 32 stations respectively were considered usable).

### 6.4.3 Data analyses

## Exploratory analyses of survey data

Table 6.4.6 provides the estimates for the burrow density and abundance for each survey. Figure 6.4 .5 shows the range of densities experienced across the ground with the higher densities occurring in the centre of the survey area and diminishing towards the perimeter. In 2009 a significant numbers of burrow systems were still apparent at the southern edge of the survey grid which suggests that the survey area may need extending further south in 2010 to better delineate the ground. Figure 6.4.6 shows the frequency and range of different densities occurring over the ground on each survey and compares the overall estimate with those from other survey areas. The mean burrow density falls at the lower end of the range of densities seen on the other grounds assessed at this working group.

The limited number of stations available on the 2007 survey and the poor quality of the data processed preclude its use in this assessment. Despite there still being some uncertainties about the spatial limits of the stock and the characteristics of the ground in this fishery the estimates still provide a good measure of abundance. In light of SGSURV and WKNEPH (2009) the data will still require further analysis and a further survey to qualify the precision of these estimates. These results therefore are presented as provisional.

The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (2009). A number of potential biases were highlighted including those due to edge effects; species burrow mis-identification and burrow occupancy. Using the same process adopted at WKNEPH, a cumulative bias correction factor for this FU was predicted to be 1.2 for FU14 (see Annex) which means the TV survey is likely to overestimate Nephrops abundance by $20 \%$.

### 6.4.4 MSY considerations

As discussed in Section 2.2 no dynamic population model is fitted to the data so no estimates of spawning stock and recruitment were available to determine Fmsy. In response to the recommendations of WKFRAME (2010), the Bell/Dobby combined sexlength cohort analysis (LCA) model (WKNEPH, 2009) was adapted to determine Harvest Rates associated with fishing at $\mathrm{F}_{35 \% \mathrm{~s} \text { SR }}$ as well as $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ (WGNSSK, 2010). These F estimates could be used as a proxy for Fmsy. Catch-length data were available for 2006, 2007, 2008 and 2009. The apparent change to the catch-length distribution for males in 2009 and the concerns about the estimate for the discarded component precluded using this data in the model. The scale and effect of any underrecording of landings pre 2006 was not known. The reference period 2006 to 2008 was selected as reflecting a recent period of relative stability for this stock despite the record landings in 2007. Figure 6.4 .7 shows the estimated selection pattern and residuals and YPR curves, from the model. Figure 6.4 .8 shows the spawner-per-recruit plot from the same model.

The results of the model in the text table below show the F multipliers required to achieve the potential $\mathrm{F}_{\mathrm{MSY}}$ proxies; the harvest rates that correspond to those multiplers and the resulting level of spawner-per-recruit as a percentage of the virgin level.

|  |  | Fbar 20-40 mm |  | Harvest <br> Rates | SPR |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Female | Male |  | Female | Male |
| $\mathrm{F}_{0.1}$ | Combined | 0.10 | 0.14 | 9.8\% | 44.6\% | 42.6\% |
|  | Female | 0.11 | 0.15 | 10.2\% | 43.5\% | 41.4\% |
|  | Male | 0.10 | 0.14 | 9.6\% | 45.3\% | 43.3\% |
| $\mathrm{F}_{35 \% \mathrm{SPR}}$ | Combined | 0.14 | 0.20 | 13.0\% | 35.9\% | 33.4\% |
|  | Female | 0.15 | 0.21 | 13.5\% | 34.7\% | 32.2\% |
|  | Male | 0.14 | 0.19 | 12.5\% | 37.1\% | 34.6\% |
| $\mathrm{F}_{\text {max }}$ | Combined | 0.20 | 0.28 | 16.4\% | 28.9\% | 26.2\% |
|  | Female | 0.21 | 0.30 | 17.4\% | 27.3\% | 24.5\% |
|  | Male | 0.19 | 0.26 | 15.8\% | 30.0\% | 27.2\% |

Following the check list presented in Section 2.2:

- Compared to other Nephrops fisheries in ICES Area VII the absolute population density of this stock is relatively low (Figure 6.4.6).
- Despite the area covered by this fishery being relatively small, the frequency distribution of the densities recorded on the two consecutive surveys 2008 and 2009 and the differences in their spatial distribution (Figure 6.4.5) suggest a degree of variation between years.
- The perception in the Irish Sea is that the growth rates in the east are similar to those in the west but the mean sizes ( mm CL ) in each fishery are markedly different, Eastern Irish Sea Nephrops being the larger.
- This fishery is highly seasonal, in effect a spring to early summer fishery, where the landings are predominantly male. Landings are around $60 \%$ male by weight and have ranged from 55 to $75 \%$ over the last 10 years.
- The annual variability of lpue for the smaller component of the catch would suggest that recruitment to this fishery, though apparently high in 2007, is quite variable. The change in discard rate and increase in the mean size of the under 35 mm component of the catch over the last two years could reflect a decline in recruitment as it does coincide with a slight decline in lpue for the same component. The rate of change in the discard rate and size could be exaggerated by poorer discard sampling in 2009, but the lpue series is more robust as the landings data are collected independently of the catch data and perceived to be well sampled.

If this decline in recruitment is real it is unclear if current levels are higher or lower than those experienced historically. Lpue overall is still high and kept high by the increasing lpue on larger males.

The two harvest ratios are $9.8 \%$ and $15.0 \%$ in 2008 and 2009 respectively
Stock density appears relatively low (Figure 6.4.6) in a highly seasonal male dominant fishery so sperm limitation could be a concern if this fishery is overexploited. To limit the potential of overfishing the males to meet a female MSY only the combined sex FMSY and male proxies are considered appropriate. Guidelines suggest a combined
sex Fmsy proxy as appropriate as long as the the virgin spawner recruit for males for that proxy does not fall below $20 \%$ (Section 2.2). At all levels of F the spawner-perrecruit is well above $20 \%$ for males.

The relatively stable trends in the mean size and the increasing lpue of the larger length group for males; the relative stability in the sex ratio and long-term decline in apparent effort suggests the current levels of effort have been sustainable. In this instance, therefore, the default proxy of $\mathrm{F}_{35 \% \mathrm{Spr}}$ is considered appropriate as it will still deliver long-term yield with a low probability of recruitment overfishing.

The relatively low densities for the two years of the UWTV survey cannot be assumed to be indicative of the population potential. The time-series for the TV surveys is too short to base a $B_{\text {trigger }}$ on one of the estimated abundances.

No $\mathrm{B}_{\text {trigger }}$ is available and a proxy for $\mathrm{F}_{\text {MSY }}$ as $\mathrm{F}_{35 \% \text { SPR combined sex }}$ is advised.

### 6.4.5 Short-term projections

A landings príiction for 2011 was made for FU14 using the approach agreed at the Benchmark Workshop (WKNEPH, 2009). The table below shows landings predicted at a range of harvest ratios including those equivalent to fishing at Fmsy proxies for the fishery as well as Fcurrent, Only the Harvest Rates associated with the male and combined sex Fmsy proxies are identified in the table as they are considered more appropriate for this stock (see below). All $\mathrm{F}_{\text {msy }}$ proxy harvest rate values are considered preliminary and may be modified following further data exploration and analysis.

| Harvest Rate |  | Implied fishery |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Survey Index (Millions) | Retained number (Millions) | Landings (tonnes) |
|  | 0\% | 214.6 | 0 | 0.00 |
|  | 2\% | " | 4 | 92.40 |
|  | 4\% | " | 9 | 184.80 |
|  | 6\% | " | 13 | 277.20 |
|  | 8\% | " | 17 | 369.59 |
| F0.1Male | 9.62\% | " | 21 | 444.58 |
| Fo.1Comb | 9.81\% | " | 21 | 453.38 |
|  | 10\% | " | 21 | 461.99 |
|  | 12\% | " | 26 | 554.39 |
| $\mathrm{F}_{35 \% \text { Male }}$ | 12.50\% | " | 27 | 577.31 |
| $\mathrm{F}_{35 \% \mathrm{Comb}}$ | 13.00\% | " | 28 | 600.44 |
|  | 14\% | " | 30 | 646.79 |
| $\mathrm{F}_{\text {current }}$ | 15.02\% | " | 32 | 694.00 |
| $\mathrm{F}_{\text {maxMale }}$ | 15.79\% | " | 34 | 729.62 |
|  | 16\% | " | 34 | 739.19 |
| $F_{\max C o m b}$ | 16.4\% | " | 35 | 756.35 |
|  |  |  | Basis |  |
| Landings Mean Weight (kg) |  | 0.0289 | Sampling 2006-2008 |  |
| Survey Bias |  | 1.2 | As per WKNEPH 2009 (See Annex) |  |
| Survey Numbers (Millions) |  | 257.5 | UWTV Survey 2009 |  |
| Proportion of removals retained by the fishery |  | 0.79 | Sampling 2006-2008 |  |

As Nephrops are advised on the basis of Harvest Rates and Fcurrent is above Fmsy a transition calculation will be required. Assuming linearity between Harvest Rate and F this will be:

$$
H R_{2011}=\left(\left(H R_{2009} \times 0.8\right)+\left(H R_{35 \% \text { SPR combined sex }} \times 0.2\right)\right)
$$

### 6.4.6 Biological reference points

Suggestions for proxies of biological reference points are shown in the catch option table.

### 6.4.7 Management plans

A number of cod recovery measures have been introduced since 2000 to conserve and promote recovery of Irish Sea cod stocks. These include a closure of the western Irish Sea cod spawning grounds from mid February to end of April since 2000, with a later extension to the eastern Irish Sea. Despite a partial derogation for Nephrops vessels during the closed period the distribution of effort on Nephrops has been affected by this management plan. There have also been various decommissioning schemes to reduce fishing effort. A $25 \%$ effort reduction on cod is in hand along with technical measures to reduce cod bycatch.

### 6.4.8 Uncertainties and bias in assessment and forecast

There are several key uncertainties and bias sources in the TV survey estimates (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to en-sure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007, WKNEPHBID 2008, SGNEPS 2009). Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that was more accurate but no more precise (WKNEPH 2009).

The cumulative bias estimates for FU14 are based on expert opinion. However these were based on experience on other Nephrops grounds and the limited survey experience on these grounds could make these less reliable in the long term. The precision of these estimates cannot yet be characterised. Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates.

### 6.4.9 Quality of assessment

The length composition and sex ratio of catches have generally been well sampled over the last ten years by E \&W. However the variability in the discard rate and selectivity within this fishery would suggest that sampling needs to be carried out at a higher level to improve on discard estimates. Discard sampling in 2009 was not as intense as in previous years which affected the confidence in the interpretation of the final inflection in the most recent trends in discard rates and catch information.

Confidence in the trends in lpue and landings has improved in the last four years.
Underwater TV surveys have been conducted annually for this stock since 2007. The quality of the data and the limited number of valid stations from the first survey limits the number of useable surveys to 2008 and 2009.

### 6.4.10 Management considerations

ICES and STECF have repeatedly advised that management should be at a smaller scale than the ICES Division level. Management at the Functional Unit level could confer controls to ensure effort and catch were in line with the scale of the resource.

In view of uncertainties about historical catch statistics interpretation of trends in lpue prior to 2006 should be treated with caution. Recent catch, effort and historical trends in size still offer some reference to the status of the stock. The reliability of landings statistics has improved and effort appears to be relatively stable although evidence would suggest it has become more targeted. There are no explicit recruitment indices.

Good catch rates of all size ranges with no significant increases in effort and no marked changes in sex ratio suggests that this stock appears to be sustaining current levels of effort. . There appears to be a slight increase in average length for the males $<35 \mathrm{~mm}$ category and a slight downturn in the lpue series from 2007 onwards which could indicate a slight decline in recruitment.

The new UWTV survey data allows for the provision of catch options and also to adopt the MSY approach. The UWTV surveys are conducted annually and a benchmarked process has been adopted. Over the last four years this stock has only been assessed biannually. These data provide the opportunity to reassess this stock more reliably on an annual basis.

Table 6.4.1. ICES Division Vlla, North of $53^{\circ}$ N: Total Nephrops landings (tonnes) by Functional Unit plus Other rectangles, 2000-2009.

| Year | FU14 | FUl5 | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 567 | 8370 | 1 | 8938 |
| 2001 | 532 | 7441 | 3 | 7976 |
| 2002 | 577 | 6793 | 1 | 7371 |
| 2003 | 376 | 7052 | 3 | 7431 |
| 2004 | 472 | 7267 | 25 | 7764 |
| 2005 | 570 | 6554 | 103 | 7227 |
| 2006 | 628 | 7561 | 52 | 8241 |
| 2007 | 959 | 8491 | 83 | 9533 |
| 2008 | 676 | 10508 | 122 | 11306 |
| $2009^{*}$ | 694 | 9198 | 57 | 9949 |

[^11]Table 6.4.2. Irish Sea East (FU14): Landings (tonnes) by country, 2000-2009.

| Year | Rep. of Ireland | UK | Other countries ** | Total |
| :---: | :---: | :---: | :---: | :---: |
| 2000 | 114 | 451 | 2 | 567 |
| 2001 | 26 | 506 | 0 | 532 |
| 2002 | 203 | 373 | 1 | 577 |
| 2003 | 69 | 306 | 1 | 376 |
| 2004 | 62 | 409 | 1 | 472 |
| 2005 | 34 | 536 | 0 | 570 |
| 2006 | 34 | 594 | 0 | 628 |
| 2007 | 86 | 873 | 0 | 959 |
| 2008 | 29 | 646 | 0 | 676 |
| $2009^{*}$ | 16 | 678 | 0 | 694 |

* provisional
** Other countries includes Belgium and Isle of Man

Table 6.4.3. Irish Sea East (FU14): Effort ('000 hours trawling) and lpue (kg/hour trawling) of Nephrops directed voyages by UK trawlers, 2000-2009.

| Year | Effort | LPUE |  |
| :---: | :---: | :---: | :---: |
| 2000 | 17.9 | 21.2 |  |
| 2001 | 20.3 | 20.7 |  |
| 2002 | 14.7 | 20.1 |  |
| 2003 | 14.1 | 16.7 |  |
| 2004 | 12.1 | 27.5 |  |
| 2005 | 13.8 | 28.5 |  |
| 2006 | 13.1 | 29.6 |  |
| 2007 | 15.8 | 39.7 |  |
| 2008 | 13.8 | 35.3 |  |
| $2009^{*}$ | 12.0 | 38.8 |  |

* provisional

Table 6.4.4. Irish Sea East (FU14): Effort ('000 hours trawling) and lpue ( $\mathrm{kg} / \mathrm{hour}$ trawling) of Nephrops directed voyages by Republic of Ireland trawlers, 2000-2009.

|  | Effort |  | LPUE |
| :---: | :---: | :---: | :---: |
|  |  | 2.5 | 43.6 |
| 2000 | 0.5 | 43.9 |  |
| 2002 | 3.3 | 57.1 |  |
| 2003 | 1.1 | 37.6 |  |
| 2004 | 1.4 | 42.8 |  |
| 2005 | 0.8 | 40.6 |  |
| 2006 | 0.7 | 53.7 |  |
| 2007 | 1.7 | 49.3 |  |
| 2008 | 0.6 | 41.6 |  |
| $2009^{*}$ | 0.4 | 40.1 |  |

* provisional

Table 6.4.5. Irish Sea East (FU14): Mean sizes (mm CL) of male and female Nephrops from UK vessels landing in England and Wales, 2000-2009.

|  | Catch |  | Landings |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Males | Females |  | Males | Females |
| 2000 | 29.2 | 28.3 | 33.7 | 32.3 |  |
| 2001 | 31.6 | 29.2 | 34.2 | 32.5 |  |
| 2002 | 32.0 | 29.2 | 35.1 | 32.0 |  |
| 2003 | 36.4 | 30.7 | 38.4 | 34.5 |  |
| 2004 | 32.2 | 29.4 | 35.2 | 33.1 |  |
| 2005 | 32.8 | 29.9 | 34.6 | 32.3 |  |
| 2006 | 33.8 | 31.4 | 36.1 | 32.6 |  |
| 2007 | 31.7 | 30.0 | 33.5 | 32.1 |  |
| 2008 | 33.0 | 30.0 | 34.0 | 31.4 |  |
| $2009^{*}$ | 34.5 | 31.3 | 34.6 | 31.8 |  |

* provisional

Table 6.4.6. Irish Sea East (FU14): Results from NI/ROI/E\&W collaborative UWTV surveys of Nephrops grounds in 2007-2009. Not corrected for bias.

| Year | Area | No. stations | Non Zero stations | Mean density | Abundance | 95\% <br> confidence interval |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{km}^{2}$ |  |  | burrows/m ${ }^{2}$ | millions | millions |
| 2007 | 1043 | 20 | 18 | 0.38 | 393 | 202 |
| 2008 | 1043 | 35 | 31 | 0.36 | 377 | 87 |
| 2009 | 1043 | 32 | 29 | 0.25 | 258 | 77 |



Figure 6.4.1. Irish Sea East (FU14). Long-term trends in landings, effort, lpues and mean sizes of Nephrops.


Figure 6.4.2. Irish Sea East (FU14). Landings, effort and lpues by quarter and sex from UK Nephrops directed trawlers.



Figure 6.4.4. Irish Sea East (FU14): Length frequency distributions of male and female landings and catch, 1997-2009.


Figure 6.4.5. Irish Sea East (FU14): Station distribution and relative burrow density, from August TV surveys 2007-2009.
a)



2008
2009
b)


Figure 6.4.6. Irish Sea East (FU14): (a) Frequency distribution of densities on UWTV survey. (b) Estimated burrow density compared with most recent density estimates from surveys carried out on other Nephrops populations. The bars indicate the range of density estimates observed over the time-series.


Figure 6.4.7. Irish Sea East (FU14): (a) Length compositions (b) Selection patterns (c) Residuals and (d) Yield-per-recruit curve and from combined sex-length cohort analysis (LCA) model (reference period:2006 to 2008). Male fine line; Female dotted line; Combined Sex bold line.


Figure 6.4.8. Irish Sea East (FU14): Spawner-per-recruit plot from combined sex-length cohort analysis (LCA) model (reference period: 2006 to 2008). Male red; Female black; Combined Sex green.

## Introduction

## Stock description and management units

A TAC is in place for ICES Areas VII which does not correspond to the assessment units. As Nephrops are limited to muddy habitats the distribution of suitable sediment defines the species distribution and the stocks are therefore assessed as six separate Functional Units (Figure 6.1.1). There are also some smaller catches from areas outside these Functional Units.

| Section | FU no. | Name | ICES <br> Divisions | Statistical rectangles |
| :--- | :---: | :--- | :--- | :--- |
| 5.4 .34 .1 | 14 | Irish Sea East | VIIa | 35-38E6; 38E5 |
| 5.4 .34 .2 | 15 | Irish Sea West | VIIa | 36E3; 35-37 E4-E5; 38E4 |
| 5.4 .34 .3 | 16 | Porcupine Bank | VIIb,c,j,k | 31-36 D5-D6; 32-35 D7-D8 |
| 5.4 .34 .4 | 17 | Aran Grounds | VIIb | 34-35 D9-E0 |
| 5.4 .34 .5 | 19 | Ireland SW and SE coast | VIIa,g,j | 31-33 D9-E0; 31E1; 32E1-E2; 33E2-E3 |
| 5.4 .34 .6 | $20-22$ | Celtic Sea | VIIg,h | 28-30 E1; 28-31 E2; 30-32 E3; 31 E4 |



Figure 6.1. Nephrops Functional Units in Subarea VII. The TAC covers all of Subarea VII. The stock area FU15 is shaded yellow.

### 6.5 Irish Sea West, FU15

### 6.5.1 General

## Type of assessment in 2010

Although the assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the general process defined by WKNEPH (2009) described in the Stock Annex new MSY target reference points were explored.

## The fishery in 2009

The Nephrops fishery in the Irish Sea west is economically the most important in ICES Division VIIa and is mainly prosecuted by vessels from UK (Northern Ireland) and Ireland. Over 100 vessels from Northern Ireland and 56 Irish vessels reporting landings from this area in 2009. A decommissioning programme was in operation in Ireland during 2007 and 2008. 14 vessels active in the FU15 fishery were decommissioned. These vessels accounted for approximately $28 \%$ of the Irish landings in 20072008.

Working Group landings from FU 15 are presented in Table 6.5.1 and Figure 6.5.1. Total declared international Nephrops landings reported from FU15 in 2009 was 9198 t and was the second highest since 1999. Ireland's landings were 2343 t and were lower than in 2008. This was accompanied by a significant reduction in effort. UK vessels landed 6855 t which was the second highest and Northern Ireland landings contributed to over $95 \%$ of this figure.

Although there has been a steady reduction in effort by the UK fleet accompanied by a migration of some vessels to the North Sea there was a slight effort increase in 2008 followed by a drop in 2009 (Table 6.5.2). Ireland's effort showed a marked reduction in 2009 (Table 6.5.3) and a rise in lpue in 2009 to a record high whilst Northern Ireland lpue decreased slightly. The mean sizes of Nephrops in the catches of both the Northern Ireland and Ireland fisheries have fluctuated without obvious trend for many years (Table 6.5.4-6.5.5, Figure 6.5.1).

Discarding of undersized and unwanted Nephrops occurs in this fishery with Northern Ireland discarding 35.9\% and Ireland 40.5\% of the catch by number in 2009 (Table 6.5.6).

Further general information on the fishery can be found in the Stock Annex.

## ICES advice applicable to 2009

## "Single-stock exploitation boundaries

Exploitation boundaries in relation to precautionary limits
The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to 2007. This corresponds to landings of no more than 8500 tonnes for the Western Irish Sea stock."

## ICES advice applicable to 2010

## "Single-stock exploitation boundaries

## June

ICES advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Rate for Nephrops fisheries should not exceed F0.1. This corresponds to landings of no more than $5465 t$ for the western Irish Sea stock.

## November

ICES advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Rate for Nephrops fisheries should not exceed F0.1. This corresponds to landings of no more than $5892 t$ for the western Irish Sea stock."

### 6.5.2 Data

An overview of the data provided and used by the WG is shown in Table 2.1. Commercial size composition data for landings and discards were provided by Northern Ireland and Ireland. Other biological data used in the assessment were as listed in the Stock Annex compiled by the Benchmark meeting WKNEPH (2009).

## Surveys

Since 2003 Ireland and Northern Ireland have jointly carried out underwater television surveys of the main Nephrops grounds in the western Irish Sea. These surveys were based on a randomised fixed grid design. The methods used during the surveys were similar to those employed for UWTV surveys of other Nephrops stocks and were as agreed by WKNEPHTV, WKNEPBID, SGNEPS and WKNEPH. An average of 145 valid stations was covered by the two surveys combined and the data were raised to a stock area of around $5340 \times 10^{-6} \mathrm{~km}^{2}$ as detailed in Table 6.5.7. Details of the survey methodology are available in WKNEPHTV.

From the time-series available, the mean density estimates calculated by the UWTV survey appeared to be very high in the initial years 2003 and 2004. The seabed in the Western Irish Sea has very high densities of Nephrops and other burrowing megafauna which makes counting a specialist task. SGNEPS 2009 reported that a random selection of $30 \%$ of the UWTV stations from 2003 and 2004 were subjected to verification in order to check for drift in burrow identification criteria over time. It concluded there was a drift at high density stations and recommended that all nonzero UWTV stations in initial years be verified by experienced counters. This was completed during 2009 and the results, which demonstrate a decrease from the initial abundance estimates in 2003 and 2004, are presented in Working Document 7. More recent abundance estimates have not been revised.

In addition to UWTV surveys Northern Ireland have completed spring (April) and summer (August) Nephrops trawl surveys since 1994 and provide data on catch rates, size composition and biological data from fixed stations in the western Irish Sea as detailed in the Stock Annex (Figure 6.5.2). The summer trawl survey catch rates correlate somewhat with UWTV survey abundance estimates (Figure 6.5.5). The longer time-series of the trawl survey shows that catch rates in the last few years (2005-2009) are close to the mean of the series when UWTV burrow abundances were in the range
of 5-6 billion burrows. Mean carapace length-by-sex has remained stable over the time-series (Figure 6.5.2).

### 6.5.3 Historical stock development

The UWTV survey method assumes that the width of the viewed transects is the entire lower edge of the TV screen on which the burrows are counted. This can be calculated from the TV camera parameters and the position of the camera in relation to the seabed. The new camera and sledge employed in 2008 for the Northern Ireland leg of the survey was used again in 2009 and gave good resolution footage. Figure 6.5.3 shows the distribution of stations sampled in 2009 which was a slightly offset grid from those sampled in 2008. Although in early surveys the distance over the ground was estimated from the vessel position alone, recent surveys use USBL positioning technology located on the sledge to give a more accurate estimate of UWTV track. The field of view of the camera at the bottom of the screen was estimated at 75 cm assuming that the sledge was flat on the seabed (i.e. no sinking). This field of view was confirmed for the majority of UWTV tows using lasers.

A re-working of the UWTV survey abundances for 2003 and 2004 were presented to the meeting and burrow abundance time-series were krigged as described in WD 7. These data along with other meta-data are shown in Table 6.5.7 and Figure 6.5.4. Figure 6.5.6 is a contour plot of the krigged density estimates for FU15 over the period 2003-2009.

The use of the UWTV surveys for the provision of Nephrops management advice was extensively reviewed by WKNEPH (ICES, 2009) and potential biases were highlighted including those due to edge effects; species burrow mis-identification and burrow occupancy. A cumulative bias correction factor estimated for FU15 was 1.14 which means the TV survey is likely to overestimate Nephrops abundance by $14 \%$.

### 6.5.4 MSY explorations

As discussed in Section 2.2 no dynamic population model is fitted to the data so no estimates of spawning-stock and recruitment were available to determine FMSY. In response to the recommendations of WKFRAME (2010), the Bell/Dobby combined sex-length cohort analysis (LCA) model used to determine Harvest Rates associated with fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ at WKNEPH (2009) was adapted to also output estimates of $\mathrm{F}_{35 \% \mathrm{Spr}}$. These F estimates could be used as a proxy for $\mathrm{F}_{\text {msy }}$. The underwater TV survey is presented as the best available information on the FU15 Nephrops stock and provides a fishery independent estimate of Nephrops abundance. Catch-length data were available for Ireland and Northern Ireland for 2008 and 2009 and were used in an SLCA model along with the biological parameter described in the Stock Annex. For other stocks three years of length data were used in the analysis but in this case there was a gap in sampling in 2006 and 2007. YPR curves and other plots generated by the model are shown in Figure 6.5.8. The F multipliers required to achieve the various Fmsy proxies are shown in the text table below along with the harvest rates that correspond to those multipliers.

|  |  | Fbar 20-40 mm | Harvest Rate | \% Virgin Spawner per Recruit |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
|  | Female | Male |  | Female |  | Male |
| $\mathrm{F}_{0.1}$ | Comb | 0.14 | 0.16 | $10.6 \%$ | $40.6 \%$ | $44.0 \%$ |
| $\mathrm{~F}_{0.1}$ | Female | 0.13 | 0.16 | $10.2 \%$ | $41.7 \%$ | $45.1 \%$ |
| $\mathrm{~F}_{0.1}$ | Male | 0.14 | 0.17 | $11.0 \%$ | $39.5 \%$ | $42.9 \%$ |
| $\mathrm{~F}_{35 \%}$ | Comb | 0.18 | 0.22 | $13.4 \%$ | $33.1 \%$ | $36.2 \%$ |
| $\mathrm{~F}_{35 \%}$ | Female | 0.17 | 0.20 | $12.7 \%$ | $34.7 \%$ | $37.9 \%$ |
| $\mathrm{~F}_{35 \%}$ | Male | 0.19 | 0.23 | $14.1 \%$ | $31.6 \%$ | $34.6 \%$ |
| $\mathrm{~F}_{\max }$ | Comb | 0.24 | 0.29 | $17.1 \%$ | $25.5 \%$ | $28.0 \%$ |
| $\mathrm{~F}_{\max }$ | Female | 0.24 | 0.29 | $17.1 \%$ | $25.5 \%$ | $28.0 \%$ |
| $\mathrm{~F}_{\max }$ | Male | 0.24 | 0.29 | $17.1 \%$ | $25.5 \%$ | $28.0 \%$ |

WGCSE took into account the following considerations:

- Compared to other Nephrops fisheries in the ICES area the population density of FU15 is the highest of all stocks $>\sim 1 / \mathrm{m} 2$ (Figure 6.5.9). These high densities are observed throughout time and space. The high observed density implies intense competition for space and food on the seabed and that sperm limitation is not likely to be a problem.
- The seven year time-series of UWTV data for FU15 and the 2009 survey shows the stock is relatively stable. Trawl survey cpue since 1994 indicates that abundance has been at high levels over the last seven years (assuming constant survey catachability).
- The growth rate of Nephrops in this stock is known to be slow and they exhibit a relatively small size of maturity (McQuaid et al.). There appears to be little change is the size composition in catches despite over 40 years of intensive fishing (Lordan, 2010, WD2).
- This fishery occurs throughout the year and does not exhibit major inter annual changes seasonal pattern. Landings have fluctuated around 9000 t for over the 35 years.
- Larval production studies show that over $440 \times 10^{9}$ larvae were produced in 1995 (Briggs et al., 2002). This $>70$ times more larvae produced annual than current stock size estimates. The high larval production is coupled with a strong retention mechanism and depositional environment due to the western Irish Sea gyre ensures continued good recruitment (Hill et al., 1994).
- The harvest rate in recent years is thought to have been above $\mathrm{F}_{\max }$ (note: harvest rates prior to 2007 are lower bounds as landings may have been under reported) with no apparent affect on the stock (Figure 6.5.10).

The WG concluded that a combined sex $F_{\text {max }}$ was a suitable $F_{\text {msy }}$ proxy for this stock. This corresponds to a harvest rate of $\mathbf{1 7 . 1} \%$.

### 6.5.5 Short-term projections

A landings prediction for 2011 was made for FU15 using the approach agreed at the Benchmark Workshop (WKNEPH ICES, 2009). Catch option table inputs are given in (Table 6.5.7) and summarised below.

Basis: Bias corrected survey index $(2009)=4.6$ billion, Mean weights $(15.1 \mathrm{~g})$ in landings (2008-2009) and retention factors based (73\%) on 2008-2009 sampling.

The landings corresponding to various fishing mortality reference points are shown in the table below.

|  |  | Landings <br> 2011 <br> (tonnes) |
| :--- | ---: | :--- |
| Rationale | Harvest ratio | 8,724 |
| MSY framework | $17.1 \%$ | 9,104 |
| MSY transition | $17.8 \%$ | 9,199 |
| F 2009 | $18.0 \%$ | 5,612 |
| Fo.1 $^{(c o m b i n e d) ~}$ | $11.0 \%$ | 6,832 |
| $\mathrm{~F}_{35 \%}$ (combined) | $13.4 \%$ | 8,724 |
| $\mathrm{~F}_{\max }$ (combined) | $17.1 \%$ |  |

## MSY transition scheme

Assuming the WG recommendation that $\mathrm{F}_{\mathrm{msy}}=\mathrm{F}_{\max }=\mathrm{HR}$ of $17.1 \%$ is accepted the following transition scheme applies: The ICES MSY framework implies the harvest ratio should be reduced ( $0.8 \times$ harvest ratio $\left(\mathrm{F}_{2009}\right)+0.2 \times$ harvest ratio $\left(\mathrm{F}_{\text {msy }}\right)$ to $17.8 \%$ resulting in landings of 9104 t in 2011.

### 6.5.6 Biological reference points

The cpue data from the trawl surveys was scaled to the UWTV index to provide a $B_{\text {trigger }}$ approximation based on the mean of the five lowest survey catch rates in the time-series (Figure 6.5.5). Harvest ratios equating to a range of fishing mortalities including $\mathrm{F}_{0.1}, \mathrm{~F}_{35} \%$ and $\mathrm{F}_{\max }$ are provided above. These calculations assumed that the TV survey has a knife-edge selectivity at 17 mm and that the supplied length frequencies represented the population in equilibrium. The WG concluded that a combined sex $\mathrm{F}_{\max }$ was a suitable $\mathrm{F}_{\text {msy }}$ proxy for this stock. This corresponds to a harvest rate of 17.1\%.

### 6.5.7 Management plans

A number of cod recovery measures have been introduced since 2000 to promote recovery of Irish Sea cod stocks. These include a closure of the western Irish Sea cod spawning grounds from mid February to end of April since 2000, with a later extension to the eastern Irish Sea closure. Despite a partial derogation for Nephrops vessels during the closed period the distribution of effort on Nephrops has been affected by this management plan. There have also been decommissioning schemes to reduce fishing effort.

### 6.5.8 Uncertainties in the assessment and forecast

There are several key uncertainties and bias sources in the method used here (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007, WKNEPHBID 2008, SGNEPS 2009). These have lead to a revision in the historical time-series of survey abundance estimates for FU15. These new estimates are thought to be of higher quality. Ultimately there still remains a degree of subjectivity in the production of UWTV abundance estimates (Marrs et al., 1996).

Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that was more accurate but no more precise (WKNEPH 2009). The survey estimates themselves are very precisely estimated (CVs 3-5\%) given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU15 are largely based on expert opinion (see Stock Annex). The precision of these bias corrections cannot yet be characterised but is likely to be higher than that observed in the survey.

In the provision of catch options based on the absolute survey estimates additional uncertain-ties related to mean weight in the landings and the discard rates also arise. For FU15 deterministic estimates of the mean weight in the landings and discard rates for 2008 and 2009 are used although there is some variability in these over time.

There is a gap of 16 months between the survey and the start of the year for which the assessment is used to set management levels. It is assumed that the stock is in equilibrium during this period (i.e. recruitment and growth balance mortality) although this is rarely the case. The effect of this assumption on realised harvest rates has not been investigated but remains a key uncertainty.

The quality of landings data has improved since 2007 with the implementation of sales notes and buyers and sellers legislation. Prior to that there were concerns that landings were underreported. The harvest ratio may be under estimated prior to 2007.

### 6.5.9 Management considerations

The FU15 Nephrops fishery first developed in the late 1950s. Since then it has sustained landings of around 9000 t for more than 35 years. Fishing effort in the past has been very high but has declined somewhat in recent years. The environment in the Western Irish Sea is very suitable for Nephrops with a large mud patch and gyre which retains the larvae over the mud patch thus ensuring good recruitment. The ground can be characterised as an area of very high densities of small Nephrops. All available information indicates that size structure of catches appears to have changed little since the fishery first began. Nevertheless the current analyses suggest that that stock is over-fished in relation to $\mathrm{F}_{\max }$ which is considered and an appropriate $\mathrm{F}_{\text {msy }}$ proxy. Higher long-term yields could be achieved by reducing fishing mortality to $\mathrm{F}_{\text {msy. }}$

The Nephrops trawl fisheries take bycatches of other species, especially juvenile whiting but also cod. Catches of these species should be reduced to as low as possible a level because of the poor status of these stocks.

The cod long-term plan was introduced in 2009 (EC 1342/2008). Annual effort in Nephrops trawl fisheries (Effort group TR2 OTB 70-99mm) in Division VIIa has been reduced by $25 \%$ in 2009 and a further $25 \%$ in 2010 and is expected to be very restrictive. The implementation of the plan is expected to cause large changes in fishing patterns. Vessels may also start using more selective gears to reduce cod catches to less than 1 or $1.5 \%$ of total catch. In 2009, three Irish vessels began using "Swedish grids" in the fishery and significantly reduced bycatches of cod, whiting and haddock (STECF 01-2010).

ICES has repeatedly advised that management should be at a smaller scale than the ICES Subarea VII. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort are at the same scale as the resource.

### 6.5.10 References

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STECF 01-2010 33rd Plenary meeting report of the Scientific, Technical and Economic Committee for Fisheries (PLEN-10-01) plenary meeting, 26-30 April 2010, Norwich. Edited by John Casey and Hendrik Dörner.

Table 6.5.1. Irish Sea West (FU15): Landings (tonnes) by country, 2000-2009.

| Year | Rep. of <br> Ireland | Isle of <br> Man | Other <br> countries |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 3,433 | 0 | 4937 | 0 | 8370 |  |  |
| 2001 | 2,689 | 3 | 4749 | 0 | 7441 |  |  |
| 2002 | 2,291 | 1 | 4501 | 0 | 6793 |  |  |
| 2003 | 2,709 | 4 | 4352 | 0 | 7065 |  |  |
| 2004 | 2,786 | 13 | 4470 | 1 | 7270 |  |  |
| 2005 | 2,133 | 0 | 4420 | 0 | 6554 |  |  |
| 2006 | 2,051 | 1 | 5508 | 1 | 7561 |  |  |
| 2007 | 2,767 | 0 | 5724 | 0 | 8491 |  |  |
| 2008 | 3,132 | 50 | 7323 | 2 | 10508 |  |  |
| $2009 *$ | 2,343 | 1 | 6855 | 0 | 9198 |  |  |

* provisional

Table 6.5.2. Irish Sea West (FU15):Landings (tonnes), effort ('000 hours trawling), and lpue ( $\mathrm{kg} /$ hour trawling) of Northern Ireland Nephrops trawlers, 2000-2009.

| Year | Landings | Effort | LPUE |
| :---: | :---: | ---: | :---: |
| 2000 | 4758 | 168.7 | 28.2 |
| 2001 | 4587 | 163.7 | 28.0 |
| 2002 | 4495 | 130.8 | 34.4 |
| 2003 | 4146 | 136.1 | 29.0 |
| 2004 | 4273 | 144.3 | 29.6 |
| 2005 | 4235 | 138.4 | 30.6 |
| 2006 | 5356 | 144.1 | 37.2 |
| 2007 | 5512 | 126.9 | 43.4 |
| 2008 | 7056 | 141.4 | 49.9 |
| $2009 *$ | 6487 | 134.7 | 48.2 |

* provisional

Table 6.5.3. Irish Sea West (FU15): Catches and landings (tonnes), effort ('000 hours trawling), cpue and lpue (kg/hour trawling) Republic of Ireland Nephrops Directed Trawlers 2000-2009.

| Year | Effort | Landings | LPUE |
| :---: | :---: | :---: | :---: |
| 2000 | 61.1 | 3160 | 51.7 |
| 2001 | 52.4 | 2475 | 47.2 |
| 2002 | 49.0 | 2238 | 45.7 |
| 2003 | 45.4 | 2680 | 59.1 |
| 2004 | 51.5 | 2535 | 49.3 |
| 2005 | 48.6 | 2062 | 42.4 |
| 2006 | 50.6 | 1959 | 38.7 |
| 2007 | 48.0 | 2578 | 53.7 |
| 2008 | 47.1 | 3076 | 65.3 |
| $2009^{*}$ | 34.0 | 2290 | 67.3 |

* provisional

Table 6.5.4. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Northern Ireland catches, landings and discards, 2000-2009.

| Year | Catches |  | Landings |  | Discards |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females | Males | Females |
| 2000 | 27.7 | 24.5 | 29.4 | 26.3 | 22.5 | 22.6 |
| 2001 | 25.7 | 23.6 | 26.1 | 24.4 | 21.7 | 21.2 |
| 2002 | 26.7 | 24.1 | 26.7 | 24.9 | 21.8 | 21.7 |
| 2003 | na | na | na | na | na | na |
| 2004 | na | na | na | na | na | na |
| 2005 | na | na | na | na | na | na |
| 2006 | na | na | na | na | na | na |
| 2007 | na | na | na | na | na | na |
| 2008 | 25.9 | 24.6 | 26.9 | 25.5 | 21.4 | 21.5 |
| $2009^{*}$ | 27.6 | 25.1 | 29.3 | 26.5 | 23.6 | 23.2 |

* provisional na = not available

Table 6.5.5. Irish Sea West (FU15): Mean sizes (mm CL) of male and female Nephrops in Republic of Ireland catches, landings and discards, 2000-2009.

| Year | Catches |  | Landings |  | Discards |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females | Males | Females |
| 2000 | 29.1 | 27.1 | 32.2 | 29.7 | 24.3 | 24.0 |
| 2001 | 26.7 | 24.8 | 28.6 | 27.0 | 23.0 | 22.2 |
| 2002 | 28.9 | 25.4 | 30.2 | 27.8 | 24.6 | 23.6 |
| 2003 | 27.7 | 24.9 | 29.7 | 26.9 | 24.0 | 23.1 |
| 2004 | 28.1 | 26.1 | 29.7 | 27.8 | 23.9 | 23.7 |
| 2005 | 28.5 | 26.8 | 30.1 | 29.1 | 23.9 | 23.2 |
| 2006 | 27.7 | 25.5 | 29.5 | 27.1 | 23.8 | 23.1 |
| 2007 | 27.7 | 25.4 | 29.8 | 27.9 | 24.0 | 23.3 |
| 2008 | 27.4 | 24.6 | 28.9 | 26.6 | 22.0 | 21.4 |
| $2009^{*}$ | 28.5 | 26.3 | 30.5 | 29.2 | 24.3 | 23.4 |

* provisional

Table 6.5.6. Irish Sea West (FU15): Proportion discarded by weight and number from FU15. (note a $10 \%$ survivorship of discards is assumed in HR and forecast calculations).

| Year | Discards By Weight | Discards by number |
| :---: | :---: | :---: |
| 1986 | 0.14 | 0.27 |
| 1987 | 0.14 | 0.24 |
| 1988 | 0.07 | 0.15 |
| 1989 | 0.08 | 0.16 |
| 1990 | 0.03 | 0.07 |
| 1991 | 0.03 | 0.08 |
| 1992 | 0.13 | 0.22 |
| 1993 | 0.17 | 0.29 |
| 1994 | 0.13 | 0.25 |
| 1995 | 0.18 | 0.32 |
| 1996 | 0.14 | 0.27 |
| 1997 | 0.12 | 0.23 |
| 1998 | 0.15 | 0.27 |
| 1999 | 0.21 | 0.35 |
| 2000 | 0.22 | 0.36 |
| 2001 | 0.22 | 0.36 |
| 2002 | 0.20 | 0.31 |
| 2003 | 0.27 | 0.42 |
| 2004 | 0.22 | 0.34 |
| 2005 | 0.18 | 0.31 |
| 2006 | 0.23 | 0.36 |
| 2007 | 0.28 | 0.42 |
| 2008 | 0.12 | 0.20 |
| 2009 | 0.24 | 0.37 |
| Max | 0.28 | 0.42 |
| Min | 0.03 | 0.07 |
| Average | 0.16 | 0.28 |

Table 6.5.7. Irish Sea West (FU15): Results from NI/ROI collaborative UWTV surveys of Nephrops grounds in 2003-2009 (before and after revalidation).

Survey estimates as used provided last year. Shade values have been revised (see WD 7).

| Ground | Year | Number <br> of stations | Mean <br> Density (No./M²) | Domain <br> Area (km²) | Unrevised estimate (billions) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Western Irish Sea | 2003 | 160 | 1.25 | 5292 | 7.0 | 1\% |
|  | 2004 | 147 | 1.52 | 5302 | 8.5 | 2\% |
|  | 2005 | 141 | 1.08 | 5288 | 6.0 | 3\% |
|  | 2006 | 138 | 1.07 | 5429 | 5.9 | 4\% |
|  | 2007 | 148 | 1.00 | 5452 | 5.6 | 3\% |
|  | 2008 | 141 | 0.88 | 5287 | 4.9 | 3\% |
|  | 2009 | 142 | 0.95 | 5267 | 5.3 | 3\% |

Revalidated survey data

| Ground | Year | Number of stations | Mean <br> Density (No./M²) | Domain Area (km²) | Revised Estimate (billions) | CV on Burrow estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Western Irish Sea | 2003 | 160 | 1.12 | 5295 | 6.3 | 3\% |
|  | 2004 | 147 | 1.13 | 5310 | 6.3 | 3\% |
|  | 2005 | 141 | 1.16 | 5281 | 6.5 | 4\% |
|  | 2006 | 138 | 1.10 | 5194 | 6.2 | 4\% |
|  | 2007 | 148 | 1.06 | 5285 | 5.9 | 3\% |
|  | 2008 | 141 | 0.88 | 5287 | 4.9 | 3\% |
|  | 2009 | 142 | 0.95 | 5267 | 5.3 | 3\% |

Table 6.5.8. Irish Sea West (FU15): Catch option table inputs. Data used for 2011 catch prediction are shaded.

| Year | Landin gs in Numbe r (millio ns) | Discar <br> ds in <br> Numbe <br> r <br> (millio <br> ns) | Remov als in Numbe <br> r <br> (millio <br> ns) | Prop <br> Remov als <br> Retaine <br> d | Adjust <br> ed <br> Survey <br> (billion <br> s) | Harve <br> st <br> Ratio | Landin gs (t) | Discar ds ( t ) | Mean <br> Weigh <br> t in <br> landin <br> gs (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 740 | 268 | 981 |  |  |  | 9,978 | 1,680 |  |
| 1987 | 774 | 242 | 992 |  |  |  | 9,753 | 1,608 |  |
| 1988 | 576 | 104 | 669 |  |  |  | 8,586 | 639 |  |
| 1989 | 644 | 121 | 753 |  |  |  | 8,147 | 673 |  |
| 1990 | 678 | 53 | 726 |  |  |  | 8,308 | 276 |  |
| 1991 | 792 | 65 | 850 |  |  |  | 9,566 | 345 |  |
| 1992 | 525 | 151 | 661 |  |  |  | 7,547 | 1,079 |  |
| 1993 | 679 | 275 | 926 |  |  |  | 8,102 | 1,622 |  |
| 1994 | 619 | 203 | 801 |  |  |  | 7,606 | 1,185 |  |
| 1995 | 554 | 260 | 787 |  |  |  | 7,796 | 1,724 |  |
| 1996 | 469 | 170 | 622 |  |  |  | 7,247 | 1,202 |  |
| 1997 | 731 | 214 | 924 |  |  |  | 9,971 | 1,330 |  |
| 1998 | 616 | 229 | 822 |  |  |  | 9,128 | 1,560 |  |
| 1999 | 710 | 388 | 1060 |  |  |  | 10,780 | 2,913 |  |
| 2000 | 533 | 298 | 801 |  |  |  | 8,370 | 2,293 |  |
| 2001 | 573 | 315 | 857 |  |  |  | 7,438 | 2,112 |  |
| 2002 | 491 | 223 | 692 |  |  |  | 6,792 | 1,732 |  |
| 2003 | 404 | 291 | 666 | 0.61 | 5.48 | 0.12 | 7,052 | 2,659 | 17.5 |
| 2004 | 416 | 218 | 612 | 0.68 | 5.55 | 0.11 | 7,267 | 1,993 | 17.5 |
| 2005 | 346 | 157 | 488 | 0.71 | 5.67 | 0.09 | 6,530 | 1,412 | 18.9 |
| 2006 | 467 | 261 | 701 | 0.67 | 5.40 | 0.13 | 7,534 | 2,285 | 16.1 |
| 2007 | 511 | 375 | 848 | 0.60 | 5.15 | 0.16 | 8,424 | 3,246 | 16.5 |
| 2008 | 755 | 191 | 927 | 0.81 | 4.29 | 0.22 | 10,478 | 1,421 | 13.9 |
| 2009 | 566 | 337 | 870 | 0.65 | 4.62 | 0.19 | 9,199 | 2,949 | 16.2 |
| Max | 792 | 388 | 1060 | 0.81 | 5.67 | 0.22 | 10,780 | 3,246 | 18.9 |
| Min | 346 | 53 | 488 | 0.60 | 4.29 | 0.09 | 6,530 | 276 | 13.9 |
| Average | 590 | 225 | 793 | 0.68 | 5.17 | 0.15 | 8,400 | 1,664 | 16.7 |
| Avg. 08-09 |  |  |  | 0.73 |  |  |  |  | 15.06 |



Figure 6.5.1. Irish Sea West (FU15): Long-term trends in landings, effort, cpues and/or lpues, and mean sizes of Nephrops.



Figure 6.5.2. Irish Sea West (FU15): Nephrops catches, sex ratio mean size from NI trawl surveys.


+ 2009_plan
$\square$ Ireland_UK

Figure 6.5.3. Irish Sea West (FU15): UWTV Stations for 2009 survey.


Figure 6.5.4. Irish Sea West (FU15): UWTV index revised and unrevised estimates.


Figure 6.5.5. Irish Sea West (FU15): Revised UWTV index and scaled trawl survey. Cpue along with Btrigger based upon mean of 5 lowest trawl survey values

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Figure 6.5.6. Irish Sea West (FU15): Contour plots of the krigged density estimates for the Irish Sea from 2003-2009.


Figure 6.5.7. Irish Sea West (FU15): Burrow density distributions 2003-2009.


Figure 6.5.8. Irish Sea West (FU15): Outputs from LCA model.


Figure 6.5.9. Irish Sea West (FU15): Estimated burrow density compared with most recent density estimates from surveys carried out on other Nephrops populations.


FU 15: Harvest rate


Figure 6.5.10. Irish Sea West (FU15): Stock summary plot of landings (tonnes), UWTV abundance and harvest rate (ratio).

### 6.6 Whiting in VIIa

## Type of assessment

This year single fleet SURBA runs were carried out for two of the main surveys assessing this stock, the NIGFS March and NIGFS October to provide trends in the stock. Overall it is clear that the stock is in a state of decline. Landings have decreased, and have been at low levels in recent years $\nless 100$ t). The survey results indicate a decline in SSB to low levels in recent years. Total mortality has been variable over the time-series.

ICES advice applicable to 2009
The Single Stock Exploitation Boundary advised by ICES for 2009 was as follows:

- Exploitation boundaries in relation to precautionary limits.

On the basis of the stock status ICES advises that catches of whiting in 2009 should be the lowest possible.

## ICES advice applicable to 2010

The Single Stock Exploitation Boundary advised by ICES for 2010 was as follows:

- Exploitation boundaries in relation to precautionary limits.

On the basis of the stock status ICES advises that catches of whiting in 2010 should be the lowest possible.

### 6.6.1 General

## Stock description and management units

The stock and the management unit are both ICES Division VIIa (Irish Sea).


## Management applicable to 2009 and 2010

The minimum landing size of whiting is 27 cm . The 2010 TAC for whiting VIIa has been reduced from 209 t to 157 t . This TAC has not been considered restrictive, with officially reported VIIa landings totalling t in 2009.

TAC 2009

| Species: Whiting <br> Merlangius merlangus |  | Zone: | VIIa <br> (WHG/07A.) |
| :---: | :---: | :---: | :---: |
| Belgium | 1 |  |  |
| France | 7 |  |  |
| Ireland | 120 |  |  |
| The Netherlands | 0 |  |  |
| United Kingdom | 81 |  |  |
| EC | 209 |  |  |
| TAC | 209 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TAC 2010

| Species:Whiting <br> Merlangius merlangus | Zone:VIla <br> (WHG/07A.) |  |
| :--- | :--- | :--- | :--- |
| Belgium | 0 |  |
| France | 5 |  |
| Ireland | 91 |  |
| The Netherlands | 0 |  |
| United Kingdom | 61 | Analytical TAC |
| EU | 157 |  |
| TAC | 157 |  |

## Fishery in 2009

ICES officially reported landings for Division VIIa and landings as used by the Working Group are given in Table 6.6.1. In recent years the values provided to the WG are very similar to officially reported landings. In 2009 international landings provided to the Working Group have increased by $25 \%$ to those of 2008, although actual numbers remain extremely low, 100 t .

The Irish Sea whiting stock is primarily caught by otter trawlers and to a lesser extent, Scottish seines, beam trawls and gillnets. Otter trawlers utilize two main mesh size ranges, $70-89 \mathrm{~mm}$ and $100-119 \mathrm{~mm}$. Effort of trawlers utilizing the larger mesh range, traditionally targeting whitefish (cod, haddock, whiting) has seen a large declined since 2003, partially as a result of effort management restrictions. The smaller range however has remained relatively stable. The primary target species of this smaller mesh range is Nephrops from which whiting is discarded at a high rate.

The closure of the western Irish Sea to whitefish fishing from mid-February to the end of April, designed to protect cod, was continued in 2009 but is unlikely to have affected whiting catches which are mainly bycatch in the derogated Nephrops fishery. Nephrops vessels can obtain a derogation to fish in certain sections of the closed area, providing they fit separator panels to their nets to allow escape of cod and other fish. The Irish and UK NI Nephrops fishery shows a peak in activity in summer months, after the reopening of the Irish Sea cod box.

In late 2009, a number of Irish vessels operating within the Irish Sea Nephrops fishery incorporated a Swedish grid into otter trawls, as part of the cod long-term management plan. It is expected that this will reduce the whiting catches of these vessels by $\sim 60 \%$ in weight. Furthermore, a small number of vessels began utilizing an inclined separator panel expected to reduce whiting catch by $\sim 75 \%$ in weight (STECF, 2010).

For a fourth successive year, Irish East Coast Nephrops vessels have moved away from their traditional Irish Sea grounds to the Smalls grounds (FU20; VIIg) which is not controlled by effort limitation and generally better prices are obtained for their catch.

During 2008 Ireland introduced a further decommissioning scheme with the aim of removing 11140 GT from the fleet register. This was targeted at vessels over 10 years and $>18 \mathrm{~m}$. Of the decommissioned vessels 29 operated within the Irish Sea, primarily targeting Nephrops landing into east, and to a lesser extent south coast ports.

### 6.6.2 Data

An overview of the data provided and used by the WG is shown in Table 2.1 in the WGCSE Report.

## Fishery landings

Table 6.6.1 gives the nominal landings of VIIa whiting as reported by each country to ICES. The officially reported landings have declined since 1996. Landings remained at a very low level in 2009, although show an increase of $25 \%$ to 2008 , Working Group estimates of catch available since 1980 are illustrated in Figure 6.6.1 and indicate the declining trend since the start of the time-series. Minor revisions were made to last year's Working Group estimate of landings ( $\sim 0.5 \mathrm{t}$ ).

There is evidence that officially reported landings of whiting in the past (especially around the mid 1990s) have been inaccurate due to misreporting. Landings data have previously been partially corrected for by using sample-based estimates of landings at a number of Irish Sea ports. Due to the low level of landings recently, this has not been carried out since 2003.

The introduction of UK and Irish legislation requiring registration of fish buyers and sellers may mean that the reported landings from 2006 onwards are more representative of actual landings.

Sampling and raising methods previously used are described in the Stock Annex for VIIa whiting. Methods for estimating quantities and composition of landings are described in the Stock Annex (Section B1.1).

Landings, discards and total catch numbers and weights-at-age for the period 1980 to 2002 as estimated by WGNSDS 2002 are given in Tables 6.6 .3 to 6.6.8. The proportion of the total catch comprising of discards from the Nephrops fleets increased over time for ages 1 and above (Table 6.6.9), although this will also reflect trends in catch of vessels not sampled for discards. While the proportion of discarded fish has increased it is largely due to the decline in abundance of marketable sized whiting
( $>27 \mathrm{~cm}$ ) and the total volume over time has declined as shown in Table 6.6.10. Mean weights-at-age for landings and discards are presented in Figure 6.6.3.

Since 2003 it has not been possible to construct catch numbers-at-age for this stock. This is due to a number of factors including low levels of landings, leading to low sampling levels, in addition to restricted access to some ports in some years.

## Discards data

Discarding of whiting is high within the Irish Sea. The onboard observer trips carried out in 2009 by UK (E\&W), UK (NI) and Ireland, showed negligible fish were retained on board, while high numbers of small fish were discarded. Raised discards from the main national fleets landing whiting show over 40 million whiting, 1500 t in weight, were discarded in 2009. This focused on the two youngest ages, and to a lesser extent age 2 . In some years up to age 4 fish are discarded. The following discard data were available for this stock:

- Discard numbers-at-age from 1980-2002 estimated from the NI Nephrops fishery and raised to the International Fleet (from the NI self sampling scheme).
- Discard numbers-at-age from the Irish Otter Trawl Fleet from 1996-2009, including length frequency data.
- Discard Length Frequencies for the UK (E and W) fleet, 2004-2009, raised to trip.
- Discard numbers-at-age for the NI fleet for 1997-2001, and 2006, 2007 and 2009, raised to trip, including length frequency data from the NI observer scheme.

Methods for estimating quantities and composition of discards from UK (NI) and Irish Nephrops trawlers are described in the Stock Annex Section B.1.2. Irish otter trawl fleet discard estimates (1996-2009), raised according to the methods described in Borges et al., 2005 were available to the Working Group (Table 6.6.11).

Mean weights-at-age for the Irish otter trawl fleet are also presented (Figure 6.6.4(b)).
The length frequency of discards of national sampled fleets in 2009 is given in Figure 6.6.5.

## Biological data

The derivation of these parameters and variables is described in the Stock Annex 6.6.

## Survey data used in assessment

Table 6.6.2 describes the survey data made available to the Working Group. Slight revisions to the UK (E\&W)-BTS-3Q survey time-series were made for the 2009 Working Group.

Figure 6.6.2 provides a comparison of mean catch weights of whiting from the eastern and western Irish Sea for UK NIGFS-Mar groundfish surveys from 1992 to March 2010 indicating low level catch rates since 2003. There is some indication within the 2009 NIGFS-Oct survey of a good recruitment, the signal for this far weaker in the following March survey.

Further information on whiting distribution is detailed in the results of Fisheries Science Partnership surveys of Irish Sea round fish stocks (www.cefas.co.uk/fsp,WD10).

WGNSDS 2006 also provides information on the distribution of whiting less than MLS in the Irish Sea up to 2006.

Survey-series for whiting provided to the Working Group are further described in the Stock Annex for VIIa whiting (SectionB.3).

## Commercial cpue

Commercial catch and effort series data available to the Working Group are described in the Stock Annex for VIIa whiting (Section B.4). Although effort data were provided for the UK (E\&W) and Ireland, it was decided not to include this data in the Report as it was considered not to be indicative of lpue trends due to the low levels of landings and changes in discard practices.

### 6.6.3 Historical stock development

No assessment was carried out for this stock in 2008 or 2009. The last assessment for this stock was a survey based assessment in 2007.

Catch-at-age data was not updated and commercial catch data was not explored in 2009.

## Data screening

The general methodology is outlined in Section 2.

## Final update assessment

Single fleet survey based runs were carried out on the NIGFS-Mar and NIGFS-Oct surveys using SURBA (version 2.2). Default values of 1 were used for both catchability and Lambda settings.

Log-mean standardised indices and scatter plots of log-index at age for the NIGFSMarch are presented in Figures 6.6.6(a) and Figure 6.6.7(a), respectively. Both plots indicate poor internal consistency within the survey. The survey appears to track the 1991 year class but examination of the internal consistency via the scatter plots indicates poor correlation between age classes. Corresponding figures for the NIGFS-Oct are plotted in Figures 6.6.6(b) and 6.6.7(b) for the UK Northern Ireland October groundfish Survey. There is some indication of tracking for the 1991, 1994 and 1995 year class but scatterplots at age are noisy and do not show strong positive correlations.

Catch curves for the NIGFS-Mar and NIGFS-Oct survey are plotted in Figure 6.6.8(a) and (b). Both surveys show a steep decline in log-numbers at age over time.

Empirical SSB estimates are presented in Figure 6.6.9 for the NIGFS March and the NIGFS October surveys. Both NIGFS surveys show SSB to be leveling out at highly reduced levels since around 2005/2006.

Figure 6.6 .10 shows the residual plots by age for the NIGFS March survey, the model fits well for age one but for older ages residuals are quite noisy, especially in the latter part of the time-series. Stock summary for the NIGFS March is shown in Figure 6.6.11. The temporal F trend is variable in later years with the current year being comparatively low, there are no extreme age or cohort effects. The plot of empirical SSB with model fit (bottom, centre) shows good fit for most years. Figure 6.6.12 shows the retrospective summary plot for the NIGFS March survey. SSB is declining since 2002, and has reached low levels in most recent years; there is no apparent re-
trospective pattern. F shows an increasing trend over the time-series, although it appears to have temporarily declined in 2008. Recruitment is also variable but estimated to be been good in 2006 and 2008. There is no strong retrospective pattern for recruitment but there are noisy periods 1995-2000 and 2004-2008.

Residual plots by age for the NIGFS October survey are shown in Figure 6.6.13. Residuals are quite noisy for all ages apart from age 0 . Figure 6.6 .14 shows the stock summary plot for the NIGFS October. The temporal F trend is variable throughout the time-series, particularly within the last 3 years. There appears to be an age effect for age 3 for this survey but no strong cohort effects. The plot of empirical SSB versus model estimates shows deteriorating fit for the latter part of the time-series. Retrospective patterns for the summary plots (Figure 6.6.15) show a variable F trend over the time-series. SSB has been declining since 2003 with a slight halt in 2008. Recruitment appears to have been good in 2006 and 2008. No retrospective bias is evident in F, SSB or recruitment.

## The state of the stock

The decline in fishery landings to under 1000 t since 2000 has been interpreted in all assessment models as a collapse in biomass, despite the absence of an analytical assessment. Generally, trends in biomass have been declining in recent years. Recruitment appears to have been good in 2006 and 2008. However the long-term trends of recruitment for this stock are difficult to interpret given the uncertainty in discard estimates for younger ages.

### 6.6.4 Short-term predictions

### 6.6.5 Medium-term projection

There is no analytical assessment for this stock.

### 6.6.6 Maximum sustainable yield evaluation

High discarding, low landings and poor sampling has lead to uncertain catch data in recent years. This data does not support the evaluation or estimation of $\mathrm{F}_{\text {msy }}$. However it is likely that current $F$ is above $F_{\text {msy }}$.

### 6.6.7 Biological reference points

Precautionary approach reference points
No precautionary reference points have been defined for this stock.

### 6.6.8 Management plans

No management plan has been agreed or proposed.

### 6.6.9 Uncertainties and bias in assessment and forecast

There is no analytical assessment for this stock.

### 6.6.10 Recommendations for next benchmark assessment

Before a benchmark can be recommended, it is first necessary to construct international catch numbers/weights-at-length and age for the main fleets engaged in the fishery since 2003. Effort data for the main fleets engaged in whiting VIIa fisheries are
required to provide a time-series of trends in commercial lpue. None of these issues will be resolved in the short term and a benchmark assessment of this stock in the near future is unlikely.

### 6.6.11 Management considerations

Technical measures applied to this stock include a minimum landing size 2 ( cm ) and minimum mesh sizes applicable to the mixed demersal fisheries. These measures are set depending on areas and years by several regulations.

Whiting are caught within a number of different fisheries as a non-target species, primarily within demersal otter trawl fisheries. Significant decline of the mixed gadoid directed fishery has occurred within the Irish Sea to minimal levels. Bycatches also occur within flatfish and ray beam trawl fisheries.

Discarding of this stock is a major consideration and efforts should be made to reduce catches of undersized fish through technical considerations. In late 2009, a number of Irish vessels operating within the Irish Sea Nephrops fishery incorporated a Swedish grid into otter trawls, as part of the cod long-term management plan. It is expected that this will reduce the whiting catches of these vessels by $\sim 60 \%$ in weight. Furthermore, a small number of vessels began utilizing an inclined separator panel expected to reduce whiting catch by $\sim 75 \%$ in weight (STECF, 2010). Implementation of such measures should be actively encouraged.

Effort limitations are in force within the Irish Sea as a result of the cod long-term management plan. Although vessels catching whiting will be affected by this regulation at present it is not believed that the effort limitations will prove beneficial to the whiting stock.

Whiting has a low market value, which is likely to contribute to discarding rates.

### 6.6.12 References

STECF 01-2010. 33rd Plenary meeting report of the Scientific, Technical and Economic Committee for Fisheries (PLEN-10-01) plenary meeting, 26-30 April 2010, Norwich. Edited by John Casey and Hendrik Dörner.

Table 6.6.1. Nominal catch ( t ) of whiting in Division VIIa, 1988-2009, as officially reported to ICES and Working Group. Discard estimates available until 2001.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 90 | 92 | 142 | 53 | 78 | 50 | 80 | 92 | 80 | 47 | 52 |
| France | 1,063 | 533 | 528 | 611 | 509 | 255 | 163 | 169 | 78 | 86 | 81 |
| Ireland | 4,394 | 3,871 | 2,000 | 2,200 | 2,100 | 1,440 | 1,418 | 1,840 | 1,773 | 1,119 | 1,260 |
| Netherlands |  |  |  |  |  |  |  |  | 17 | 14 | 7 |
| UK(Engl. \& Wales) ${ }^{\text {a }}$ | 1,202 | 6,652 | 5,202 | 4,250 | 4,089 | 3,859 | 3,724 | 3,125 | 3,557 | 3,152 | 1,900 |
| Spain |  |  |  |  |  |  |  |  |  |  |  |
| UK (Isle of Man) | 15 | 26 | 75 | 74 | 44 | 55 | 44 | 41 | 28 | 24 | 33 |
| UK (N.Ireland) | 4,621 |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 107 | 154 | 236 | 223 | 274 | 318 | 208 | 198 | 48 | 30 | 22 |
| UK |  |  |  |  |  |  |  |  |  |  |  |
| Total human consumption | 11,492 | 11,328 | 8,183 | 7,411 | 7,094 | 5,977 | 5,637 | 5,465 | 5,581 | 4,472 | 3,355 |
| Estimated Nephrops fishery discards used by the $\mathrm{WG}^{\text {b }}$ | $1,611$ | 2,103 | 2,444 | 2,598 | 4,203 | 2,707 | 1,173 | 2,151 | 3,631 | 1,928 | 1,304 |
| Working Group Estimates | 11,856 | 13,408 | 10,656 | 9,946 | 12,791 | 9,230 | 7,936 | 7,044 | 7,966 | 4,205 | 3,533 |


| Country | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 46 | 30 | 27 | 22 | 13 | 11 | 10 | 4.2 | 3 | 2 | 2.3 |
| France | 150 | 59 | 25 | 33 | 29 | 8 | 13 | 3.7 | 3 | 2 | NA |
| Ireland | 509 | 353 | 482 | 347 | 265 | 96 | 94 | 55.3 | 187 | 68 | 67.56 |
| Netherlands | 6 | 1 |  |  |  |  |  |  |  |  |  |
| UK(Engl. \& Wales) ${ }^{\text {a }}$ | 1,229 | 670 | 506 | 284 | 130 | 82 | 47 | 21.7 | 3 | 11 | 19.9 |
| Spain |  |  |  |  | 85 |  |  |  |  |  |  |
| UK (Isle of Man) | 5 | 2 | 1 | 1 | 1 | 1 |  |  | 1 | 1 |  |
| UK (N.Ireland) |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 44 | 15 | 25 | 27 | 31 | 6 | <0.5 | <0.5 | <0.5 |  |  |
| UK |  |  |  |  |  |  |  |  |  |  |  |
| Total human consumption | 1,989 | 1,130 | 1,066 | 714 | 554 | 204 | 164 | 84.9 | 197 | 84 | 90 |
| Estimated Nephrops fishery discards used by the $\mathrm{WG}^{\text {b }}$ | 1,092 | 2,118 | 1,012 | 740 | n/a | n/a | n/a | n/a | n/a | n/a | n/a |
| Working Group Estimates | 2,762 | 2,880 | 1,745 | 1,487 | 676 | 184 | 158 | 86 | 196 | 81 | 101 |

[^12]Table 6.6.2. VIIa whiting survey data available to WGCSE 2010. Survey Titles highlighted in bold have been updated.


NIGFS West-March : Northern Ireland March Groundfish Survey - Irish Sea West - Nos. per 3 nm 19942010

| 10.21 | 0.25 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 4 |  |  |  |  |  |
| 4307 | 73 | 121 | 6 | 0 | 1994 |
| 3604 | 988 | 53 | 30 | 1 | 1995 |
| 2323 | 587 | 188 | 11 | 15 | 1996 |
| 3250 | 447 | 52 | 14 | 1 | 1997 |
| 3857 | 535 | 71 | 9 | 3 | 1998 |
| 2373 | 228 | 39 | 7 | 2 | 1999 |
| 4037 | 231 | 23 | 3 | 0 | 2000 |
| 1998 | 631 | 30 | 2 | 1 | 2001 |
| 3580 | 163 | 36 | 3 | 0 | 2002 |
| 2952 | 812 | 25 | 6 | 1 | 2003 |
| 3568 | 174 | 36 | 1 | 0 | 2004 |
| 1219 | 97 | 6 | 1 | 0 | 2005 |
| 1266 | 150 | 12 | 0 | 0 | 2006 |
| 1825 | 190 | 10 | 1 | 0 | 2007 |
| 1254 | 290 | 17 | 1 | 0 | 2008 |
| 1941 | 227 | 10 | 1 | 0 | 2009 |
| 1485 | 297 | 20 | 1 | 0 | 2010 |

NIGFS East-October : Northern Ireland October Groundfish Survey -
Irish Sea East - Nos. per 3 nm
19942009

| 1 | 10.83 | 0.88 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 5 |  |  |  |  |  |  |
| 1 | 749 | 472 | 179 | 165.0 | 29.0 | 3.0 | 1994 |
| 1 | 2515 | 259 | 178 | 41.0 | 47.0 | 9.0 | 1995 |
| 1 | 1005 | 517 | 127 | 64.0 | 15.0 | 10.0 | 1996 |
| 1 | 640 | 668 | 682 | 88.0 | 26.0 | 6.0 | 1997 |
| 1 | 1446 | 277 | 178 | 95.0 | 11.0 | 4.0 | 1998 |
| 1 | 2287 | 1388 | 260 | 102.0 | 79.0 | 3.0 | 1999 |
| 1 | 1972 | 1288 | 216 | 26.0 | 22.0 | 9.0 | 2000 |
| 1 | 2998 | 691 | 300 | 35.0 | 7.0 | 5.0 | 2001 |
| 1 | 1296 | 1285 | 349 | 76.0 | 8.5 | 2.0 | 2002 |
| 1 | 3783 | 1939 | 1104 | 155.4 | 25.0 | 3.2 | 2003 |
| 1 | 1820 | 521 | 347 | 109.1 | 7.7 | 1.7 | 2004 |
| 1 | 1247 | 865 | 296 | 17.5 | 1.9 | 0.6 | 2005 |
| 1 | 2304 | 150 | 52 | 9.0 | 2.1 | 0.0 | 2006 |
| 1 | 1094 | 827 | 165 | 18.4 | 2.9 | 3.1 | 2007 |


| 1 | 2329 | 873 | 81 | 1.3 | 0.2 | 0.0 | 2008 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 641 | 675 | 48 | 4.4 | 1.1 | 0.0 | 2009 |  |
| NIGFS East-March : Northern Ireland March Groundfish Survey - Irish |  |  | Northern Ireland March Groundfish Survey - Irish |  |  |  |  |  |
| Sea East - Nos. per 3 nm |  |  |  |  |  |  |  |  |
|  | 2010 |  |  |  |  |  |  |  |
| 1 | 10.21 | 0.25 |  |  |  |  |  |  |
| 1 | 5 |  |  |  |  |  |  |  |
| 1 | 611 | 290 | 390 | 47 | 12.0 | 1994 |  |  |
| 1 | 448 | 522 | 142 | 109 | 25.0 | 1995 |  |  |
| 1 | 1094 | 221 | 203 | 40 | 44.0 | 1996 |  |  |
| 1 | 561 | 1054 | 91 | 33 | 2.0 | 1997 |  |  |
| 1 | 409 | 903 | 522 | 32 | 11.0 | 1998 |  |  |
| 1 | 1023 | 407 | 135 | 52 | 6.0 | 1999 |  |  |
| 1 | 1481 | 524 | 229 | 35 | 4.0 | 2000 |  |  |
| 1 | 631 | 739 | 162 | 15 | 9.0 | 2001 |  |  |
| 1 | 869 | 1043 | 243 | 54 | 13.1 | 2002 |  |  |
| 1 | 1118 | 1328 | 178 | 24 | 5.7 | 2003 |  |  |
| 1 | 1026 | 302 | 69 | 4 | 1.6 | 2004 |  |  |
| 1 | 499 | 129 | 41 | 12 | 3.9 | 2005 |  |  |
| 1 | 964 | 323 | 39 | 10 | 0.7 | 2006 |  |  |
| 1 | 623 | 120 | 11 | 3 | 0 | 2007 |  |  |
| 1 | 669 | 417 | 51 | 3 | 0 | 2008 |  |  |
| 1 | 956 | 313 | 47 | 2 | 0 | 2009 |  |  |
| 1 | 671 | 357 | 24 | 2 | 2 | 2010 |  |  |
| UKE\&W-BTS : Corystes Irish Sea Beam Trawl Survey (Sept) - Prime sta tions only - Effort and numbers at age (per km towed) |  |  |  |  |  |  |  |  |
| 19882009 |  |  |  |  |  |  |  |  |
| 1 | 10.75 | 0.79 |  |  |  |  |  |  |
| 0 | 1 |  |  |  |  |  |  |  |
| 1 | 326 | 134 | 1988 |  |  |  |  |  |
| 1 | 226 | 66 | 1989 |  |  |  |  |  |
| 1 | 316 | 242 | 1990 |  |  |  |  |  |
| 1 | 494 | 74 | 1991 |  |  |  |  |  |
| 1 | 451 | 596 | 1992 |  |  |  |  |  |
| 1 | 297 | 197 | 1993 |  |  |  |  |  |
| 1 | 196 | 133 | 1994 |  |  |  |  |  |
| 1 | 1952 | 74 | 1995 |  |  |  |  |  |
| 1 | 172 | 207 | 1996 |  |  |  |  |  |
| 1 | 406 | 277 | 1997 |  |  |  |  |  |
| 1 | 905 | 186 | 1998 |  |  |  |  |  |
| 1 | 581 | 153 | 1999 |  |  |  |  |  |
| 1 | 321 | 139 | 2000 |  |  |  |  |  |
| 1 | 596 | 197 | 2001 |  |  |  |  |  |
| 1 | 283 | 103 | 2002 |  |  |  |  |  |
| 1 | 520 | 184 | 2003 |  |  |  |  |  |
| 1 | 908 | 339 | 2004 |  |  |  |  |  |
| 1 | 845 | 293 | 2005 |  |  |  |  |  |
| 1 | 1019 | 222 | 2006 |  |  |  |  |  |
| 1 | 369 | 90 | 2007 |  |  |  |  |  |
| 1 | 826 | 85 | 2008 |  |  |  |  |  |
| 1 | 397 | 385 | 2009 |  |  |  |  |  |
| NIGFS-Oct E\&W : Northern Ireland October Groundfish Survey - Irish Sea |  |  |  |  |  |  |  |  |
| East \& West - Nos. per 3 |  |  |  |  |  |  |  |  |
| 19922009 |  |  |  |  |  |  |  |  |
| 1 | 10.83 | 0.88 |  |  |  |  |  |  |
| 0 | 5 |  |  |  |  |  |  |  |
| 1 | 1454 | 995 | 96 | 26.0 | 4.0 | 0.0 | 1992 |  |
| 1 | 1554 | 425 | 300 | 27.0 | 2.0 | 0.1 | 1993 |  |
| 1 | 2450 | 686 | 133 | 123.0 | 20.0 | 2.0 | 1994 |  |
| 1 | 3199 | 483 | 163 | 30.9 | 33.6 | 6.9 | 1995 |  |
| 1 | 2628 | 605 | 124 | 50.0 | 10.8 | 6.8 | 1996 |  |
| 1 | 3219 | 655 | 504 | 63.0 | 19.0 | 4.0 | 1997 |  |
| 1 | 3601 | 414 | 164 | 70.0 | 7.9 | 3.0 | 1998 |  |
| 1 | 3945 | 1060 | 191 | 70.0 | 54.1 | 1.7 | 1999 |  |


| 1 | 2631 | 1066 | 158 | 18.0 | 15.8 | 6.1 | 2000 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 6911 | 713 | 270 | 29.0 | 4.7 | 3.1 | 2001 |  |  |  |
| 1 | 3189 | 1421 | 274 | 55.4 | 6.1 | 1.5 | 2002 |  |  |  |
| 1 | 5284 | 1831 | 901 | 111.9 | 17.4 | 2.2 | 2003 |  |  |  |
| 1 | 4892 | 712 | 276 | 78.1 | 5.3 | 1.2 | 2004 |  |  |  |
| 1 | 2583 | 684 | 219 | 14.2 | 1.5 | 0.4 | 2005 |  |  |  |
| 1 | 3045 | 157 | 43 | 7.6 | 1.6 | 0.0 | 2006 |  |  |  |
| 1 | 2638 | 1039 | 153 | 13.8 | 2.0 | 2.1 | 2007 |  |  |  |
| 1 | 5815 | 1492 | 149 | 4.1 | 0.1 | 0.0 | 2008 |  |  |  |
| 1 | 2328 | 637 | 70 | 7.6 | 1.3 | 0.0 | 2009 |  |  |  |
|  | - March | \&W : | rthe | Irela | M Mar | Gr | dfish |  | Iris | Sea |
|  | \& West | Nos. | per 3 | m |  |  |  |  |  |  |
|  | 2010 |  |  |  |  |  |  |  |  |  |
| 1 | 10.21 | 0.25 |  |  |  |  |  |  |  |  |
| 1 | 5 |  |  |  |  |  |  |  |  |  |
| 1 | 1477 | 456 | 94 | 29 | 5.0 | 0.0 | 1992 |  |  |  |
| 1 | 667 | 655 | 67 | 9 | 2.0 | 0.5 | 1993 |  |  |  |
| 1 | 1790 | 221 | 304 | 34 | 8.0 | 5.0 | 1994 |  |  |  |
| 1 | 1696 | 698 | 116 | 85 | 17.0 | 3.0 | 1995 |  |  |  |
| 1 | 1478 | 280 | 160 | 28 | 32.0 | 5.6 | 1996 |  |  |  |
| 1 | 1419 | 860 | 79 | 27 | 1.7 | 4.3 | 1997 |  |  |  |
| 1 | 1730 | 767 | 196 | 12 | 3.3 | 0.1 | 1998 |  |  |  |
| 1 | 1453 | 350 | 104 | 38 | 5.0 | 1.0 | 1999 |  |  |  |
| 1 | 2297 | 431 | 163 | 25 | 2.7 | 0.0 | 2000 |  |  |  |
| 1 | 1067 | 704 | 120 | 11 | 7 | 1.6 | 2001 |  |  |  |
| 1 | 1734 | 762 | 177 | 38 | 9 | 0.3 | 2002 |  |  |  |
| 1 | 1703 | 1163 | 129 | 18 | 4 | 0.0 | 2003 |  |  |  |
| 1 | 1837 | 261 | 59 | 3 | 1 | 0.1 | 2004 |  |  |  |
| 1 | 729 | 119 | 30 | 9 | 3 | 0.3 | 2005 |  |  |  |
| 1 | 1054 | 274 | 31 | 7 | 1 | 0.1 | 2006 |  |  |  |
| 1 | 1007 | 142 | 11 | 2 | 0.1 | 0.0 | 2007 |  |  |  |
| 1 | 856 | 376 | 40 | 3 | 0.2 | 0.0 | 2008 |  |  |  |
| 1 | 1270 | 285 | 35 | 1 | 0.1 | 0.1 | 2009 |  |  |  |
| 1 | 931 | 338 | 23 | 2 | 1.5 | 0.0 | 2010 |  |  |  |
| UKNI-MIK : Northern Ireland MIK Net Survey |  |  |  |  |  |  |  |  |  |  |
| 19942009 |  |  |  |  |  |  |  |  |  |  |
| 1 | 10.46 | 0.50 |  |  |  |  |  |  |  |  |
| 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1 | 778 | 1994 |  |  |  |  |  |  |  |  |
| 1 | 225 | 1995 |  |  |  |  |  |  |  |  |
| 1 | 397 | 1996 |  |  |  |  |  |  |  |  |
| 1 | 205 | 1997 |  |  |  |  |  |  |  |  |
| 1 | 59 | 1998 |  |  |  |  |  |  |  |  |
| 1 | 91 | 1999 |  |  |  |  |  |  |  |  |
| 1 | 40 | 2000 |  |  |  |  |  |  |  |  |
| 1 | 167 | 2001 |  |  |  |  |  |  |  |  |
| 1 | 19 | 2002 |  |  |  |  |  |  |  |  |
| 1 | 148 | 2003 |  |  |  |  |  |  |  |  |
| 1 | 101 | 2004 |  |  |  |  |  |  |  |  |
| 1 | 135 | 2005 |  |  |  |  |  |  |  |  |
| 1 | 118 | 2006 |  |  |  |  |  |  |  |  |
| 1 | 82 | 2007 |  |  |  |  |  |  |  |  |
| 1 | 99 | 2008 |  |  |  |  |  |  |  |  |
| 1 | 173 | 2009 |  |  |  |  |  |  |  |  |
| ScoGFS Spring : Scottish groundfish survey in Spring |  |  |  |  |  |  |  |  |  |  |
|  | 2006 (0.15 |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.15 | 0.21 |  |  |  |  |  |  |  |
| 1 | 8 |  |  |  |  |  |  |  |  |  |
| 1 | 11610 |  | 4051 | 1898 | 362 | 229 | 59 | 3 | 4 | 1996 |  |
| 1 | 16322 | 16200 | 2953 | 964 | 250 | 105 | 39 | 1 | 1997 |  |
| 1 | 22145 | 8187 | 3817 | 137 | 110 | 0 | 5 | 0 | 1998 |  |
| 1 | 19815 | 6642 | 1706 | 282 | 11 | 0 | 27 | 0 | 1999 |  |
| 1 | 13019 | 1662 | 169 | 71 | 36 | 6 | 0 | 0 | 2000 |  |
| 1 | 9419 | 4541 | 407 | 40 | 2 | 0 | $\bigcirc$ | 0 | 2001 |  |
| 1 | 15605 | 3060 | 430 | 34 | 1 | 0 | $\bigcirc$ | 0 | 2002 |  |


| 11 | 14798 | 5404 | 375 | 45 | $\bigcirc$ | 4 | 0 | 0 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19 | 9199 | 2219 | 583 | 27 | 1 | 0 | 0 | 0 | 2004 |
| 13 | 3783 | 899 | 200 | 56 | 3 | 0 | 0 | 0 | 2005 |
| 17 | 7317 | 1040 | 319 | 32 | 2 | 0 | 0 | 0 | 2006 |
| ScoGFS Autumn : Scottish groundfish survey |  |  |  |  |  |  |  |  |  |
| 19952005 |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.83 | 0.91 |  |  |  |  |  |  |
| $\bigcirc$ | 6 |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |
| 13 | 30094 | 8827 | 2530 | 435 | 215 | 4 | 0 |  |  |
| 11 | 18457 | 7166 | 1291 | 37 | 35 | 26 | 0 | 19 |  |
| 17 | 73309 | 7357 | 2166 | 263 | 219 | 0 | 6 | 19 |  |
| 11 | 16862 | 8677 | 503 | 242 | 25 | 12 | 0 | 20 |  |
| 10 | 0 | 140 | 133 | 13 | 0 | 0 | 0 | 20 |  |
| 13 | 30324 | 16655 | 1435 | 224 | 2 | 28 | 0 | 20 |  |
| 12 | 26671 | 7170 | 1138 | 69 | 0 | 0 | 0 | 20 |  |
| 14 | 42435 | 19333 | 3321 | 319 | 3 | 0 | 0 | 20 |  |
| 11 | 16510 | 3382 | 97 | 4 | 2 | 3 | 0 | 20 |  |
| IR-ISCSGFS : Irish Sea Celtic Sea GFS 4th Qtr - Effort min. towed |  |  |  |  |  |  |  |  |  |
| No. at age |  |  |  |  |  |  |  |  |  |
| 19972002 |  |  |  |  |  |  |  |  |  |
| 11 | 10.8 | 0.9 |  |  |  |  |  |  |  |
| 05 |  |  |  |  |  |  |  |  |  |
| 5401 | 1566 | 3330 | 793 | 154 | 23 | 12 |  |  |  |
| 10204 | 48396 | 6534 | 2249 | 170 | 15 | 0 |  |  |  |
| 11702 | 208494 | 3302 | 624 | 24 | 28 | 2 |  |  |  |
| 11289 | 97502 | 4402 | 25 | 1 | 0 | 0 |  |  |  |
| 1221 | 28881 | 29577 | 3123 | 177 | 1 | 0 |  |  |  |
| 10351 | 12112 | 10237 | 1497 | 225 | 33 | 5 |  |  |  |
| IR-Q4 IBTS: IRISH GFS RV Celtic Explorer: NUMBERS AT AGE |  |  |  |  |  |  |  |  |  |
| 20032004 |  |  |  |  |  |  |  |  |  |
| 11 | 10.89 | 0.91 |  |  |  |  |  |  |  |
| $0 \quad 5$ |  |  |  |  |  |  |  |  |  |
| 17 | 72340 | 19658 | 13391 | 1617 | 605 | 0 |  |  |  |
| 17 | 75196 | 14563 | 1293 | 147 | 5 | 2 |  |  |  |
| IR-OTB : Irish Otter trawl - Effort in h - VIIa Whiting numbers at age |  |  |  |  |  |  |  |  |  |
| 19952002 |  |  |  |  |  |  |  |  |  |
| 11 | 10 | 1 |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |
| 80314 | 6 | 437 | 206 | 261 | 21 | 1 |  |  |  |
| 64824 | 64 | 682 | 1528 | 266 | 71 | 4 |  |  |  |
| 92178 | 3 | 368 | 494 | 418 | 55 | 19 |  |  |  |
| 93533 | 20 | 395 | 838 | 117 | 27 | 30 |  |  |  |
| 110275 | 534 | 398 | 531 | 130 | 19 | 3 |  |  |  |
| 82690 | 40 | 192 | 155 | 58 | 8 | 0 |  |  |  |
| 77541 | 13 | 397 | 444 | 42 | 22 | 3 |  |  |  |
| 77863 | 21 | 173 | 383 | 88 | 8 | 8 |  |  |  |
| UKNI-Pelagic trawl : Northern Ireland Midwater trawlers - Effort in h |  |  |  |  |  |  |  |  |  |
| 19932002 |  |  |  |  |  |  |  |  |  |
| 11 | 10 | 1 |  |  |  |  |  |  |  |
| 26 |  |  |  |  |  |  |  |  |  |
| 74014 | 3174 | 1060 | 172 | 29.5 | 4.8 | 199 |  |  |  |
| 73778 | 1706 | 4340 | 574 | 72.8 | 16.2 | 199 |  |  |  |
| 52773 | 1997 | 416 | 719 | 37.9 | 7.2 | 199 |  |  |  |
| 53083 | 1432 | 2276 | 361 | 327.4 | 41.8 | 199 |  |  |  |
| 55863 | 1241 | 660 | 549 | 12.3 | 17.5 | 199 |  |  |  |
| 61153 | 438 | 423 | 98 | 45.8 | 2.7 | 199 |  |  |  |
| 72859 | 162 | 185 | 57 | 13.5 | 11.6 | 199 |  |  |  |
| 46412 | 67 | 53 | 11 | 7.9 | 1.1 | 200 |  |  |  |
| 50302 | 7 | 4 | 2 | 0.5 | 0.2 | 200 |  |  |  |


| 57754 | 189 | 316 | 90 | 11 | 15 | 2002 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UKNI-Otter trawl : Northern Ireland single-rig otter trawlers - Effort in h - No per h fished - includes discards |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| 19932002 |  |  |  |  |  |  |  |  |  |
| 11 | 10 | 1 |  |  |  |  |  |  |  |
| 06 | 6 |  |  |  |  |  |  |  |  |
| 195323 | 310308 | 9217 | 21444 | 2791 | 261 | 28 | 2 | 1993 |  |
| 191705 | 53172 | 11286 | 3957 | 9723 | 747 | 75 | 16 | 1994 |  |
| 161025 | 55228 | 10692 | 8874 | 987 | 1312 | 17 | 1 | 1995 |  |
| 154418 | 88663 | 20784 | 6748 | 4623 | 551 | 460 | 56 | 1996 |  |
| 165612 | 24344 | 12001 | 5864 | 1292 | 528 | 7 | 7 | 1997 |  |
| 149088 | 85869 | 11381 | 2368 | 1135 | 200 | 50 | 1 | 1998 |  |
| 146990 | 014625 | 3517 | 1202 | 344 | 59 | 12 | 8 | 1999 |  |
| 130117 | 74403 | 12613 | 3082 | 520 | 61 | 14 | 8 | 2000 |  |
| 131418 | 810658 | 6663 | 1833 | 228 | 64 | 13 | 10 | 2001 |  |
| 108616 | 64601 | 8586 | 1068 | 265 | 44 | 3 | 2 | 2002 |  |
| UKE\&W-Otter trawl : England/Wales Otter Trawl |  |  |  |  |  |  |  |  |  |
| 19812000 |  |  |  |  |  |  |  |  |  |
| 11 | 10 | 1 |  |  |  |  |  |  |  |
| 26 |  |  |  |  |  |  |  |  |  |
| 1079 | 906 | 766 | 162 | 103 | 4 | 1981 |  |  |  |
| 1271 | 1984 | 893 | 340 | 67 | 49 | 1982 |  |  |  |
| 88 | 685 | 1065 | 227 | 67 | 21 | 1983 |  |  |  |
| 1031 | 1395 | 439 | 475 | 80 | 29 | 1984 |  |  |  |
| 1032 | 2077 | 889 | 148 | 125 | 25 | 1985 |  |  |  |
| 90 | 2246 | 1006 | 158 | 20 | 17 | 1986 |  |  |  |
| 131 | 2206 | 1505 | 316 | 58 | 5 | 1987 |  |  |  |
| 1321 | 1885 | 827 | 161 | 30 | 6 | 1988 |  |  |  |
| 1401 | 1344 | 1201 | 234 | 40 | 10 | 1989 |  |  |  |
| 1172 | 2076 | 671 | 222 | 35 | 14 | 1990 |  |  |  |
| 1072 | 2374 | 793 | 165 | 48 | 5 | 1991 |  |  |  |
| 972 | 2072 | 1020 | 177 | 42 | 3 | 1992 |  |  |  |
| 797 | 784 | 654 | 157 | 31 | 5 | 1993 |  |  |  |
| 4311 | 110 | 454 | 91 | 15 | 3 | 1994 |  |  |  |
| 43 | 460 | 188 | 375 | 7 | 1 | 1995 | Rev | ed at NSWG | 1997 |
| 42 | 260 | 604 | 102 | 90 | 10 | 1996 |  |  |  |
| 40 | 331 | 211 | 155 | 7 | 1 | 1997 |  |  |  |
| 37 | 311 | 355 | 81 | 28 | 1 | 1998 |  |  |  |
| 231 | 194 | 175 | 46 | 11 | 8 | 1999 |  |  |  |
| 271 | 186 | 134 | 47 | 36 | 4 | 2000 |  |  |  |

Table 6.6.3. VIIa whiting International numbers-at-age ('000) for human consumption, 1980-2002 (partially corrected for misreporting). Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 14520 | 11203 | 5427 | 4886 | 18254 | 15540 | 6306 | 10149 | 6983 | 11645 |
| 2 | 21811 | 29011 | 18098 | 9943 | 12683 | 35324 | 16839 | 21563 | 25768 | 14029 |
| 3 | 6468 | 16004 | 19340 | 9100 | 5257 | 8687 | 10809 | 6968 | 6989 | 13011 |
| 4 | 2548 | 2596 | 6108 | 4530 | 2571 | 996 | 1877 | 1943 | 1513 | 3645 |
| 5 | 350 | 821 | 813 | 1165 | 1045 | 675 | 285 | 242 | 396 | 490 |
| 6+ | 621 | 339 | 400 | 321 | 402 | 372 | 270 | 111 | 197 | 177 |
| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0 | 102 | 0 | 38 | 0 | 0 | 129 | 0 | 0 | 1 |
| 1 | 9502 | 7426 | 8380 | 2742 | 3245 | 1124 | 1652 | 610 | 329 | 341 |
| 2 | 17604 | 18406 | 21907 | 21468 | 6983 | 10095 | 6162 | 4239 | 3287 | 2806 |
| 3 | 4734 | 5829 | 7959 | 7327 | 18509 | 3020 | 7432 | 2567 | 4727 | 2607 |
| 4 | 1477 | 993 | 1374 | 932 | 1801 | 4444 | 1263 | 1795 | 888 | 741 |
| 5 | 318 | 311 | 462 | 135 | 208 | 233 | 1082 | 87 | 261 | 160 |
| 6+ | 128 | 84 | 93 | 27 | 50 | 21 | 135 | 79 | 95 | 119 |
| Age | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1 | 319 | 111 | 67 |  |  |  |  |  |  |  |
| 2 | 1364 | 1189 | 748 |  |  |  |  |  |  |  |
| 3 | 1002 | 1006 | 1480 |  |  |  |  |  |  |  |
| 4 | 299 | 171 | 376 |  |  |  |  |  |  |  |
| 5 | 115 | 53 | 48 |  |  |  |  |  |  |  |
| 6+ | 15 | 20 | 41 |  |  |  |  |  |  |  |

Table 6.6.4. VIIa whiting International discard numbers-at-age (‘000), 1980-2002. Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 12786 | 9865 | 4047 | 23847 | 26394 | 12380 | 28364 | 16594 | 6922 | 17247 |
| 1 | 32318 | 24935 | 8489 | 7328 | 33900 | 26461 | 21111 | 40598 | 17958 | 20701 |
| 2 | 6888 | 9162 | 560 | 2036 | 1568 | 1859 | 1464 | 1875 | 1940 | 2476 |
| 3 | 65 | 162 | 19 | 9 | 11 | 9 | 33 | 0 | 0 | 26 |
| 4 | 26 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $6+$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 4216 | 20349 | 1497 | 12639 | 3731 | 7118 | 12732 | 8163 | 6096 | 20851 |
| 1 | 31810 | 29334 | 61451 | 13979 | 12063 | 17613 | 39647 | 25497 | 27131 | 7677 |
| 2 | 3353 | 3823 | 10404 | 17707 | 1812 | 7015 | 8168 | 5352 | 2293 | 2117 |
| 3 | 72 | 146 | 97 | 426 | 1702 | 492 | 1976 | 689 | 550 | 228 |
| 4 | 0 | 1 | 0 | 5 | 29 | 234 | 81 | 141 | 44 | 34 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| $6+$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |


| Age | 2000 | 2001 | 2002 |
| ---: | ---: | ---: | ---: |
| 0 | 7321 | 16940 | 8538 |
| 1 | 38922 | 12631 | 13412 |
| 2 | 4395 | 3150 | 1588 |
| 3 | 564 | 102 | 231 |
| 4 | 55 | 10 | 33 |
| 5 | 1 | 0 | 0 |
| $6+$ | 10 | 0 | 1 |

Table 6.6.5. VIIa whiting International catch numbers-at-age ('000) combined landings and discards, 1980-2002. Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 12786 | 9865 | 4088 | 23847 | 26394 | 12380 | 28364 | 16594 | 6922 | 17247 |
| 1 | 46838 | 36138 | 13916 | 12214 | 52154 | 42001 | 27417 | 50747 | 24941 | 32346 |
| 2 | 28699 | 38173 | 18658 | 11979 | 14251 | 37183 | 18303 | 23438 | 27708 | 16505 |
| 3 | 6533 | 16166 | 19359 | 9109 | 5268 | 8696 | 10842 | 6968 | 6989 | 13037 |
| 4 | 2574 | 2622 | 6108 | 4530 | 2571 | 996 | 1877 | 1943 | 1513 | 3645 |
| 5 | 350 | 821 | 813 | 1165 | 1045 | 675 | 285 | 242 | 396 | 490 |
| $6+$ | 621 | 339 | 400 | 321 | 402 | 372 | 270 | 111 | 197 | 177 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 4216 | 20451 | 1497 | 12677 | 3731 | 7118 | 12861 | 8163 | 6096 | 20852 |
| 1 | 41312 | 36760 | 69831 | 16721 | 15308 | 18737 | 41299 | 26107 | 27460 | 8018 |
| 2 | 20957 | 22229 | 32311 | 39175 | 8795 | 17110 | 14330 | 9591 | 5580 | 4923 |
| 3 | 4806 | 5975 | 8056 | 7753 | 20211 | 3512 | 9408 | 3256 | 5277 | 2835 |
| 4 | 1477 | 994 | 1374 | 937 | 1830 | 4678 | 1344 | 1936 | 932 | 776 |
| 5 | 318 | 311 | 462 | 135 | 208 | 233 | 1082 | 87 | 261 | 161 |
| $6+$ | 128 | 84 | 93 | 27 | 50 | 21 | 135 | 79 | 95 | 121 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |
| 0 | 7321 | 16940 | 8538 |  |  |  |  |  |  |  |
| 1 | 39242 | 12742 | 13479 |  |  |  |  |  |  |  |
| 2 | 5758 | 4338 | 2336 |  |  |  |  |  |  |  |
| 3 | 1566 | 1108 | 1711 |  |  |  |  |  |  |  |
| 4 | 354 | 181 | 409 |  |  |  |  |  |  |  |
| 5 | 115 | 53 | 48 |  |  |  |  |  |  |  |
| $6+$ | 25 | 20 | 42 |  |  |  |  |  |  |  |

Table 6.6.6. VIIa whiting International landings mean weight-at-age (kg), 1980-2002. Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.133 | 0.133 | 0.133 | 0 | 0.144 | 0 | 0.134 | 0 | 0 | 0 |
| 1 | 0.216 | 0.216 | 0.216 | 0.215 | 0.208 | 0.174 | 0.184 | 0.173 | 0.152 | 0.197 |
| 2 | 0.269 | 0.269 | 0.269 | 0.279 | 0.257 | 0.250 | 0.225 | 0.223 | 0.214 | 0.209 |
| 3 | 0.365 | 0.365 | 0.365 | 0.397 | 0.403 | 0.333 | 0.342 | 0.363 | 0.330 | 0.269 |
| 4 | 0.533 | 0.533 | 0.533 | 0.491 | 0.550 | 0.478 | 0.512 | 0.535 | 0.547 | 0.433 |
| 5 | 0.630 | 0.630 | 0.630 | 0.605 | 0.699 | 0.567 | 0.709 | 0.720 | 0.763 | 0.680 |
| $6+$ | 0.772 | 0.888 | 0.736 | 0.655 | 0.745 | 0.642 | 0.940 | 0.933 | 1.005 | 1.079 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0 | 0.115 | 0 | 0.117 | 0 | 0 | 0 | 0 | 0 | 0.120 |
| 1 | 0.198 | 0.172 | 0.160 | 0.151 | 0.169 | 0.188 | 0.196 | 0.171 | 0.169 | 0.166 |
| 2 | 0.220 | 0.210 | 0.198 | 0.186 | 0.198 | 0.219 | 0.217 | 0.219 | 0.202 | 0.218 |
| 3 | 0.313 | 0.266 | 0.274 | 0.233 | 0.227 | 0.273 | 0.244 | 0.244 | 0.240 | 0.255 |
| 4 | 0.436 | 0.352 | 0.361 | 0.332 | 0.304 | 0.334 | 0.288 | 0.296 | 0.274 | 0.328 |
| 5 | 0.676 | 0.453 | 0.513 | 0.454 | 0.378 | 0.551 | 0.365 | 0.396 | 0.350 | 0.352 |
| $6+$ | 0.800 | 0.692 | 1.007 | 0.892 | 0.496 | 1.320 | 0.415 | 0.537 | 0.421 | 0.328 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |
| 0 | 0.064 | 0 | 0 |  |  |  |  |  |  |  |
| 1 | 0.179 | 0.182 | 0.145 |  |  |  |  |  |  |  |
| 2 | 0.216 | 0.250 | 0.214 |  |  |  |  |  |  |  |
| 3 | 0.269 | 0.319 | 0.273 |  |  |  |  |  |  |  |
| 4 | 0.317 | 0.346 | 0.356 |  |  |  |  |  |  |  |
| 5 | 0.347 | 0.538 | 0.449 |  |  |  |  |  |  |  |
| $6+$ | 0.412 | 0.337 | 0.428 |  |  |  |  |  |  |  |

Table 6.6.7. VIIa whiting International discard mean weight-at-age (kg), 1980-2002. Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.034 | 0.034 | 0.029 | 0.033 | 0.024 | 0.022 | 0.023 | 0.024 | 0.021 | 0.026 |
| 1 | 0.062 | 0.062 | 0.072 | 0.101 | 0.075 | 0.080 | 0.058 | 0.078 | 0.069 | 0.063 |
| 2 | 0.125 | 0.125 | 0.125 | 0.147 | 0.130 | 0.137 | 0.126 | 0.157 | 0.114 | 0.105 |
| 3 | 0.230 | 0.230 | 0.141 | 0.245 | 0 | 0 | 0.155 | 0 | 0.449 | 0.091 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $6+$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  | 0 |
| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.034 | 0.030 | 0.014 | 0.029 | 0.029 | 0.031 | 0.026 | 0.026 | 0.017 | 0.028 |
| 1 | 0.060 | 0.051 | 0.050 | 0.050 | 0.048 | 0.055 | 0.051 | 0.041 | 0.034 | 0.038 |
| 2 | 0.113 | 0.115 | 0.110 | 0.089 | 0.123 | 0.120 | 0.111 | 0.101 | 0.090 | 0.086 |
| 3 | 0.115 | 0.130 | 0.137 | 0.143 | 0.154 | 0.153 | 0.161 | 0.141 | 0.130 | 0.147 |
| 4 | 0 | 0 | 0 | 0.175 | 0.149 | 0.179 | 0.186 | 0.170 | 0.145 | 0.237 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.218 |
| $6+$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.174 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |
| 0 | 0.024 | 0.017 | 0.016 |  |  |  |  |  |  |  |
| 1 | 0.036 | 0.034 | 0.033 |  |  |  |  |  |  |  |
| 2 | 0.100 | 0.088 | 0.082 |  |  |  |  |  |  |  |
| 3 | 0.128 | 0.119 | 0.127 |  |  |  |  |  |  | 0 |

Table 6.6.8. VIIa whiting International catch mean weight-at-age ( $\mathbf{k g}$ ) combined landings and discard, 1980-2002. Estimates have not been possible since 2003 due to low landings and resulting poor sampling.

| Age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.034 | 0.040 | 0.031 | 0.033 | 0.032 | 0.021 | 0.025 | 0.024 | 0.021 | 0.026 |
| 1 | 0.110 | 0.118 | 0.135 | 0.146 | 0.125 | 0.107 | 0.100 | 0.101 | 0.088 | 0.111 |
| 2 | 0.235 | 0.240 | 0.265 | 0.256 | 0.244 | 0.245 | 0.217 | 0.217 | 0.201 | 0.193 |
| 3 | 0.363 | 0.364 | 0.365 | 0.397 | 0.403 | 0.333 | 0.342 | 0.363 | 0.330 | 0.269 |
| 4 | 0.529 | 0.529 | 0.533 | 0.491 | 0.550 | 0.478 | 0.512 | 0.535 | 0.547 | 0.433 |
| 5 | 0.630 | 0.630 | 0.630 | 0.605 | 0.700 | 0.567 | 0.709 | 0.720 | 0.763 | 0.680 |
| $6+$ | 0.772 | 0.888 | 0.736 | 0.655 | 0.745 | 0.642 | 0.940 | 0.933 | 1.005 | 1.079 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 0.036 | 0.031 | 0.014 | 0.029 | 0.030 | 0.031 | 0.027 | 0.026 | 0.017 | 0.028 |
| 1 | 0.094 | 0.077 | 0.063 | 0.067 | 0.074 | 0.063 | 0.057 | 0.044 | 0.035 | 0.044 |
| 2 | 0.204 | 0.194 | 0.170 | 0.142 | 0.183 | 0.179 | 0.159 | 0.153 | 0.156 | 0.161 |
| 3 | 0.310 | 0.263 | 0.272 | 0.228 | 0.221 | 0.257 | 0.230 | 0.222 | 0.228 | 0.246 |
| 4 | 0.436 | 0.352 | 0.361 | 0.331 | 0.301 | 0.326 | 0.284 | 0.287 | 0.268 | 0.324 |
| 5 | 0.676 | 0.453 | 0.513 | 0.454 | 0.378 | 0.551 | 0.364 | 0.396 | 0.350 | 0.351 |
| $6+$ | 0.800 | 0.692 | 1.007 | 0.892 | 0.496 | 1.320 | 0.715 | 0.679 | 0.421 | 0.325 |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |
| 0 | 0.024 | 0.017 | 0.016 |  |  |  |  |  |  |  |
| 1 | 0.038 | 0.036 | 0.033 |  |  |  |  |  |  |  |
| 2 | 0.127 | 0.132 | 0.124 |  |  |  |  |  |  |  |
| 3 | 0.218 | 0.301 | 0.253 |  |  |  |  |  |  |  |
| 4 | 0.291 | 0.338 | 0.339 |  |  |  |  |  |  |  |
| 5 | 0.347 | 0.538 | 0.449 |  |  |  |  |  |  |  |
| $6+$ | 0.310 | 0.337 | 0.425 |  |  |  |  |  |  |  |

Table 6.6.9. VIIa whiting estimates of discard numbers-at-age from the Nephrops fleet as a proportion of total International numbers-at-age.

| Age | 0 | 1 | 2 | 3 | 4 | 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1.000 | 0.690 | 0.240 | 0.010 | 0.010 | 0 |
| 1982 | 0.990 | 0.610 | 0.030 | 0.001 | 0 | 0 |
| 1983 | 1.000 | 0.600 | 0.170 | 0.001 | 0 | 0 |
| 1984 | 1.000 | 0.650 | 0.110 | 0.002 | 0 | 0 |
| 1985 | 1.000 | 0.630 | 0.050 | 0.001 | 0 | 0 |
| 1986 | 1.000 | 0.770 | 0.080 | 0.003 | 0 | 0 |
| 1987 | 1.000 | 0.800 | 0.080 | 0 | 0 | 0 |
| 1988 | 1.000 | 0.720 | 0.070 | 0 | 0 | 0 |
| 1989 | 1.000 | 0.640 | 0.150 | 0.002 | 0 | 0 |
| 1990 | 1.000 | 0.770 | 0.160 | 0.015 | 0 | 0 |
| 1991 | 0.995 | 0.798 | 0.172 | 0.024 | 0.001 | 0 |
| 1992 | 1.000 | 0.880 | 0.322 | 0.012 | 0 | 0 |
| 1993 | 0.997 | 0.836 | 0.452 | 0.055 | 0.005 | 0 |
| 1994 | 1.000 | 0.788 | 0.206 | 0.084 | 0.016 | 0 |
| 1995 | 1.000 | 0.940 | 0.410 | 0.140 | 0.050 | 0 |
| 1996 | 0.990 | 0.960 | 0.570 | 0.210 | 0.060 | 0 |
| 1997 | 1.000 | 0.977 | 0.558 | 0.212 | 0.073 | 0 |
| 1998 | 1.000 | 0.988 | 0.411 | 0.104 | 0.047 | 0 |
| 1999 | 1.000 | 0.957 | 0.430 | 0.081 | 0.044 | 0.009 |
| 2000 | 1.000 | 0.992 | 0.763 | 0.360 | 0.154 | 0.005 |
| 2001 | 1.000 | 0.991 | 0.726 | 0.092 | 0.055 | 0 |
| 2002 | 1.000 | 0.995 | 0.680 | 0.135 | 0.081 | 0.000 |
| Mean $81-02$ | 0.999 | 0.817 | 0.311 | 0.070 | 0.027 | 0.001 |

Table 6.6.10. VIIa whiting estimated landed and discarded catch ( $\mathbf{t}$ ). Data partially corrected for misreporting.

|  | Catch (t) |  |
| :---: | :---: | :---: |
| Year | Landed | Discarded |
| 1980 | 13461 | 3324 |
| 1981 | 17646 | 2960 |
| 1982 | 17304 | 808 |
| 1983 | 10525 | 1820 |
| 1984 | 11802 | 3433 |
| 1985 | 15582 | 2654 |
| 1986 | 10300 | 2115 |
| 1987 | 10519 | 3899 |
| 1988 | 10245 | 1611 |
| 1989 | 11305 | 2103 |
| 1990 | 8212 | 2444 |
| 1991 | 7348 | 2598 |
| 1992 | 8588 | 4203 |
| 1993 | 6523 | 2707 |
| 1994 | 6763 | 1173 |
| 1995 | 4893 | 2151 |
| 1996 | 4335 | 3631 |
| 1997 | 2277 | 1928 |
| 1998 | 2229 | 1304 |
| 1999 | 1670 | 1092 |
| 2000 | 762 | 2118 |
| 2001 | 733 | 1012 |
| 2002 | 747 | 740 |
| 2003 | 401 | $n / a$ |
| Mean: | 7990 | 2253 |

Table 6.6.11. VIIa whiting discard numbers- and mean weights-at-age from the Irish otterboard trawl fleet 1996-2009.


|  | 2003 |  | 2004 |  | 2005 |  | 2006 |  | 2007 |  | 2008 |  | 2009 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Numbers ('000) | Weight (kg) | Numbers ('000) | Weight <br> (kg) | Numbers ('000) | Weight <br> (kg) | Numbers <br> ('000) | Weight <br> (kg) | Numbers ('000) | Weight <br> (kg) | Numbers ('000) | $\begin{gathered} \text { Weight } \\ (\mathrm{kg}) \end{gathered}$ | Numbers ('000) | $\begin{gathered} \text { Weight } \\ (\mathrm{kg}) \end{gathered}$ |
| 0 | 1921.76 | 0.016 | 17091.56 | 0.018 | 442.07 | 0.010 | 1534.97 | 0.016 | 5138.89 | 0.043 | 4585.77 | 0.025 | 13319.29 | 0.028 |
| 1 | 2419.56 | 0.036 | 7347.29 | 0.034 | 2531.84 | 0.035 | 1483.43 | 0.060 | 23000.16 | 0.038 | 7879.78 | 0.040 | 12913.10 | 0.036 |
| 2 | 1287.21 | 0.178 | 731.35 | 0.101 | 783.68 | 0.091 | 621.58 | 0.133 | 3282.67 | 0.095 | 1485.70 | 0.093 | 712.51 | 0.081 |
| 3 | 603.20 | 0.246 | 142.50 | 0.165 | 129.28 | 0.159 | 99.02 | 0.218 | 916.09 | 0.145 | 161.03 | 0.119 | 2.60 | 0.175 |
| 4 | 108.64 | 0.268 | 96.30 | 0.218 | 40.12 | 0.154 | 16.82 | 0.312 | 10.96 | 0.276 | 13.46 | 0.130 | 0.89 | 0.257 |
| 5 | 0.00 | 0.000 | 0.00 | 0.000 | 24.48 | 0.371 | 0.00 | 0.000 | 1.92 | 0.304 | 0.00 | 0.000 | 0.00 | 0.000 |
| 6 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 7 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 8 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 9 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 10 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 11 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 12 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 13 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| 14+ | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 | 0.00 | 0.000 |
| Total weight (t) |  | 523.6 |  | 680.3 |  | 201.3 |  | 223.2 |  | 1544.7 |  | 585.3 |  | 892.3 |
| Sampling Information | 2003 |  | 2004 |  | 2005 |  | 200 |  |  |  | 2008 |  | 200 |  |
| Number of Trips |  |  |  |  |  |  |  |  |  | 15 |  | 18 |  | 12 |
| Number of Hauls |  | 60 |  | 122 |  | 96 |  | 56 |  | 90 |  | 91 |  | 55 |



Figure 6.6.1. Whiting VIIa. Working Group estimates of landings 1980-2009. Note landings data prior to 2003 has been adjusted for misreporting and includes estimates of discards.


Figure 6.6.2. Eastern and western VIIa whiting mean catch rates in kg per 3-mile tow, for fish at and above the minimum landing size ( 27 cm ) from the UK (NI) March groundfish survey, 19922009.
a)

b)


Figure 6.6.3. VIIa whiting International mean weights-at-age in (a) landings (Human Consumption Fishery) and (b) discards, 1980-2002.
a)

b)


Figure 6.6.4. VIIa whiting discard information for the Irish commercial otterboard trawl fleet (a) numbers-at-age and (b) mean weights-at-age, 1996-2009.




Figure 6.6.5. VIIa Whiting discard length frequency by national fleets in 2009. Note due to low levels of retained catch, and hence low sampling, this data is not presented.
a)


Figure 6.6.6. Log Mean Standardized Indices for (a) NIGFS March and (b) NIGFS October by year class and year.
a)

Northern Ireland March Groundfish Survey- Irish Sea East \& West - Nos. per 3 nm: Comparative

b)

NIGFS-Oct E\&W FIXED q: Comparative scatterplots at age


Figure 6.6.7. Scatter Plots of Log index-at-age for the NIGFS March (a) and NIGFS October (b) surveys.
a)

NIGFS-March E\&W : Northern Ireland March Groundfish Survey- Irish Sea East \& West - Nos. per 3 nm: log cohort abundance

b)

NIGFS-Oct E\&W FIXED q: log cohort abundance


Figure 6.6.8. Catch Curves for NIGFS-March (a) and NIGFS-October (b) surveys.
a)


b)

NIGFS-Oct E\&W FIXED q: empirical relative SSB (unsmoothed)


Figure 6.6.9. Empirical Estimates of SSB for NIGFS March (a) and NIGFS October (b) surveys.
j-March E\&W : Northern Ireland March Groundfish Survey- Irish Sea East \& West - Nos. per 3 nm : Resi


Figure 6.6.10. Residual Plots by Age of the NIGFS March survey.
viGFS-March E\&W : Northern Ireland March Groundfish Survey- Irish Sea East \& West - Nos. per 3 nm


Figure 6.6.11. Stock Summary of the SURBA model fit for the NIGFS March survey. Empirical SSB (red dots) with model estimates of SSB (black line) are shown in bottom centre panel.

NIGFS-March E\&W : Northern Ireland March Groundfish Survey- Irish Sea East \& West - Nos. per 3 nm


Figure 6.6.12. Retrospective pattern of Single fleet SURBA run for NIGFS March survey.

NIGFS-Oct E\&W FIXED q: Residuals


Figure 6.6.13. Residual Plots by Age of the NIGFS March survey.


Figure 6.6.14. Stock Summary of the SURBA model fit for the NIGFS March survey. Empirical SSB (red dots) with model estimates of SSB (black line) are shown in bottom centre panel.


Figure 6.6.15. Retrospective pattern of Single fleet SURBA run for NIGFS March survey.

### 6.7 Plaice in Division VIla (Irish Sea)

## Type of assessment in 2010

Update assessment using the same settings as last year.

## ICES advice applicable to 2009

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects

The current fishing mortality (2007) is estimated to be 0.09, which is below the rate expected to lead to high long-term yields and low risk of stock depletion. There would be little gain to the long-term yield by increasing fishing mortalities above current levels. Fishing at Fo.1 corresponds to landings in 2009 of $1430 t$.

Exploitation boundaries in relation to precautionary limits
Fishing mortality should be kept below $F_{p a}$ (0.45). This corresponds to catches of less than 3960 t in 2009 and will maintain SSB above Bpa in 2010.

ICES advice applicable to 2010
ICES advises on the basis of high long-term yield that catches should not exceed $1627 t$ in 2010.

### 6.7.1 General

## Stock description and management units

The stock assessment area and the management unit are both Division VIIa (Irish Sea).

Management applicable in 2009 and 2010
Management of plaice in Division VIIa is by TAC and there is a minimum landing size of 27 cm in force. The agreed TACs and associated implications for plaice in Division VIIa are detailed in the tables below.

2009:

| Species: Plaice <br> Pleuronectes platessa | Zone:VIIa <br> (PLE/07A.) |  |
| :--- | ---: | :--- |
| Belgium | 37 |  |
| France | 16 |  |
| Ireland | 934 |  |
| The Netherlands | 11 |  |$\quad$| Analytical TAC |
| :--- |
| United Kingdom |
| EC |

2010:

| Species:Plaice <br> Pleuronectes platessa | Zone:VIla <br> (PLE/07A.) |  |
| :--- | ---: | :--- | :--- |
| Belgium | 42 |  |
| France | 18 |  |
| Ireland | 1063 |  |
| The Netherlands | 13 |  |
| United Kingdom | 491 |  |
| EU | 1627 | Analytical TAC |
| TAC | 1627 |  |

## The fishery in 2009

National landings data reported to ICES and Working Group estimates of total landings are given in Table 6.7.2.1.

The TAC in 2009 was 1430 tonnes and the Working Group estimate of landings in 2009 was 456 tonnes, which is a $19 \%$ decrease in landings comparable to 2008 and only $32 \%$ of the TAC in 2009. This shortfall in estimated landings relative to the TAC has occurred in previous years, increasing steadily from a $7 \%$ of TAC in 2003 to a $70 \%$ shortfall in 2008. It seems unlikely that the poor uptake of the quota is a consequence of an inability to catch sufficient quantities of plaice; rather the shortfall in the uptake of the TAC is likely due to limited consumer demand and poor value of the catch.

Landings by the Belgian, UK (E\&W) and Irish fleets comprised approximately $41 \%$, $38 \%$ and $21 \%$ respectively of total landings in 2009 . The landings of plaice are split evenly between beam trawlers (primarily Belgian vessels then Irish vessels) targeting sole and otter trawlers (primarily UK vessels then Irish vessels) fishing for whitefish or, increasingly, Nephrops.

High levels of discarding are known to occur in this fishery (see Figures 6.7.2.3 to 6.7.2.5). Previous sampling studies for discards in the Irish Sea indicate that discard-
ing of plaice is substantial (up to $80 \%$ by number) and that only a small proportion of the total catch may be retained onboard.

A general description of the fishery can be found in the Stock Annex (Stock Annex 6.6) and also in 'Other Relevant Data' section below. For general mixed fisheries advice applicable to this stock and other species taken in the same fisheries, see Section 6.1.

### 6.7.2 Data

## Landings

National landings data reported to ICES and Working Group estimates of total landings are given in Table 6.7.2.1. Landed numbers-at-age for the younger ages (ages 2 to 4) have declined more rapidly over the last two decades than landings of older fish, despite the fact that high numbers of younger fish are caught by the beam trawl survey, suggesting that the selection pattern and/or discarding behaviour of the fleets has changed over time and that the landings-at-age matrix might not be representative of the true catch (Figure 6.7.2.2). The procedures used to determine the total international landings figures are documented in the Stock Annex.

## Discards

Routine discard sampling has been conducted by the UK (E\&W) since 2000 and by Ireland since 1993. Northern Ireland has collected data from 1996 but not between 2003 and 2005, and by Belgium since 2003. Length distributions of landed and discarded fish estimates are presented for UK (E\&W) (Figure 6.7.2.3), Irish (Figure 6.7.2.4) and Belgian fleets (Figure 6.7.2.5), although Belgian data have been missing since 2007.

In 2009, observations by discard sampling trips suggest $90 \%$ discarding by number by the UK fleet and $99 \%$ by the Irish fleet.

Although these time-series of discard observations are available, they have so far not been raised to fleet level and are therefore not currently incorporated in the assessment. WKDRP has investigated the issue of raising discard samples to total catches but has not provided any clear advice on the best approach to adopt. In addition there is a considerable historical time period for which no discard sampling has taken place. Work is ongoing on the issue of raising samples and in the calculation of a historical time-series of discard data. However, raising the data remains problematic given the low sampling levels (see Working Document 8, WGNSDS 2005).

## Biological

Landings numbers-at-age are given in Table 6.7.2.5 and plotted in Figure 6.7.2.2. Weights-at-age in the landings and stock are given in Tables 6.7.2.6-6.7.2.7. The history of the derivation of the landings weights and stock weights used in this assessment is described in the Stock Annex.

Landings weights-at-age for 2009 were obtained from the quadratic fit:

$$
\mathrm{Wt}=-0.002^{*} \text { age }^{2}+0.0674^{*} \text { age }+0.0329
$$

and used a SOP correction of 0.91882 .
Landings weights-at-age calculations for this stock were problematic (large residuals about the quadratic fit) this year (2009 data) and last (2008 data) for ages greater than

12 due to the low number of sampled fish. Also UK (E\&W) and Irish values of weight-at-age show differences in weight of these older fish, which should be addressed at the next benchmark assessment.

## Surveys

All available tuning data are shown in Table 6.7.2.4. Due to inconsistencies in the available commercial tuning fleets, Irish Sea plaice assessments since 2004 have only included the UK (E\&W) beam trawl survey (UK (E\&W) BTS, September: 1989-2007) and the two UK (NI) spawning biomass indices based on groundfish surveys (UK (NI) GFS). For more information see WGNSDS 2004.

Inspection of UK (E\&W) BTS $\log$ (cpue) plots (Figures 6.7.2.6 and 6.7.2.7) indicates that the survey has fair internal consistency and suggests increases in the abundance of plaice at all ages. The biomass index calculated from the UK (E\&W) BTS during autumn (September) also indicates an upwards trend since 1991 (Figure 6.7.2.2). However, given that this survey covers only the northeastern part of the Irish Sea, it is not necessarily representative of the entire stock. In contrast, the UK (NI) groundfish surveys in March and October do cover both the northwestern and northeastern areas of the Irish Sea. Although the UK (E\&W) BTS and the UK (NI) GFS surveys show similar increases in biomass between 1992 and 2003, low biomass values were recorded between 2004 and 2007 in the autumn index of the UK (NI) surveys and between 2004 and 2009 in the spring index. Nevertheless, both autumn and spring indices reach high biomass levels in 2009 and 2010 respectively; second only to the peak in biomass in 2003 in each index suggesting that the stock is currently at high levels.

The UK (NI) ground fish surveys' strata can be disaggregated into eastern (Strata 4-7) and western (Strata 1-3) subareas, where the subareas are divided by the deep trench that runs roughly north-south to the west of the Isle of Man. The notable difference in mean biomass between spring ( 17 kg per 3 miles) and autumn ( 5 kg per 3 miles) in the western area (Strata $1-3$ ) suggests either that spawning fish migrate into the area during spring or that catchability of plaice increases during spawning. Both (March and October) time-series of the UK (NI) GFS appear dominated by change in the eastern subarea and notably this subarea includes the entire UK BTS survey area. This is particularly clear during the autumn when the correlation between the overall UK (NI) index and the eastern subarea is great, $\mathrm{R}=0.99$, which is due to the high weighting placed on stations in the larger strata i.e. in the eastern area relative to the west (Figure 6.7.2.8).

The observed increase in spawning-stock biomass by the UK (NI) GFS is pronounced in the autumn period (October samples), but not evident in the spring (March samples) when many mature fish have migrated to spawn. Indeed the autumn series for Strata 4-7 (the eastern Irish Sea) contains a statistically significant breakpoint in 1999 (supremum $(F$ statistic) $=18.6, \mathrm{p}=0.001 ; 95 \%$ confidence interval 1995-2001) after which the mean biomass ( 13.5 kg per 3 miles, 2000-2010) is double that in the early period ( 6.7 kg per 3 miles, 1992-1999).

In summary, the UK (E\&W) BTS in September and the UK (NI) GFS index in October (but not March) indicate a sustained increase in biomass in the eastern Irish Sea, but this rise does not appear to extend across the deep channel to plaice in the western Irish Sea.

The SSB of plaice in the Irish Sea is also independently estimated using the Annual Egg Production Method (AEPM):

| Year | SSB |
| :--- | :---: | :---: |
| 1995 | 9081 |
| 2000 | 13303 |
| 2006 | 11487 |
| 2008 | 12759 |

The results (revised in 2010 to ensure consistency across years, see WD 11) show substantial differences to ICES assessment values (from ICA), but they do confirm that SSB of plaice in the Irish Sea is lightly exploited. Splitting the SSB estimates from the AEPM into eastern and western Irish Sea areas also indicates that the perceived increase in plaice biomass is due to increased production in the eastern Irish Sea only in agreement with the trends noted above in BTS and GFS data (For more details see Stock Annex).

Work is currently being undertaken to supply cpue values for the Q4 western IBTS survey (UK, E\&W) for the Irish Sea area. It is anticipated that this time-series will contribute to this assessment following benchmarking of the stock.

## Commercial cpue

All available tuning data are shown in Table 6.7.2.4. Age based tuning data available for this assessment comprise 3 commercial fleets; the UK (E\&W) otter trawl fleet (UK (E\&W) OTB, 1987-2008), the UK (E\&W) beam trawl fleet (UK (E\&W) BT, 1989-2008) and the Irish otter trawl fleet (IR-OTB, 1995-2008). Due to inconsistencies in the available tuning fleets, Irish Sea plaice assessments since 2004 have omitted these indices. For more information see WGNSDS 2004.

## Other relevant data

Table 6.7.2.2 and Figure 6.7.2.1 show that effort levels have decreased between 2008 and 2009 for all fleets. Both the UK otter and beam trawl fleets are at their lowest recorded effort levels in time-series extending back to 1972 and 1978 respectively. However, anecdotal information from the UK fishing industry has suggested an abundance of plaice in Area VIIa in recent years. Belgian vessels operating in Division VII typically move in and out of the Irish Sea, depending on the season, from specifically the Bristol Channel and Celtic Sea, the Bay of Biscay and the southern North Sea.

In 2009, landings by the Belgian fleet increased by 40 tonnes relative to 2008 landings. For the UK (E\&W) , the otter trawl fleet reports the majority (approximately 99\%) of plaice landings, which are typically low in the first quarter when the fish are generally found further offshore in deeper water. The Irish fishery landings in 2009 were split mostly between otter trawlers (57\%), and beam trawlers (38\%). The beam trawl component is mostly taken as part of a mixed fishery, and some of the landings also come as bycatch from the Nephrops fishery.

Landings by the Belgian fleet in 2009 were greatest in the four quarter ( $38 \%$ ) and lowest in the first and third quarters ( $17 \%$ each). Landings by UK (E\&W) were largely taken in the second and third quarters ( $38 \%$ and $34 \%$ respectively, with the highest landings occurring in July 30.5 tonnes), and lowest during quarter one ( $9 \%$ ). Landings by the Irish fleet were greatest in the third quarter (39\%) and lowest in the first and second quarters ( $16 \%$ each).

### 6.7.3 Historical Stock development

Model: ICA
Software: FLICA. (Software versions are given in Table 6.7.3.1.)

## Model options chosen

Setting for this update stock assessment are given in the table below, as in standard for this stock the separable period has been increased by one year relative to last years assessment. The update ICA assessment follows the same procedure as in the 2009 assessment as described in Stock Annex.

## Input data types and characteristics

New data added to the update ICA assessment are the fishery landings data for 2009 and survey data for 2009 for the following surveys: UK(E\&W) BTS, UK(NI) GFS March and UK(NI) GFS October.

## Data screening

Data was screened as described in the Stock Annex. A separable VPA model was used to examine the structure of the landings numbers-at-age data before its use in update assessment. The fitted model indicates that the age structure of the recorded landings may have changed over the last decade, with increasingly negative residuals at the younger ages (ages 2 and 3 ) and increasingly positive residuals at ages $>4$. This may be a result of discarding (noted earlier) and the same effect is visible clearly in plots of standardised proportions-at-age in the landings matrix (Figure 6.7.2.2).

## Final update assessment

A summary plot for the final update ICA assessment is shown in Figure 6.7.3.3 and time-series estimates for F, SSB and recruitment are given in Table 6.7.3.4. The ICA assessment settings are shown in the following table, with changes to the previous years' settings highlighted in bold. Historical settings are given in the Stock Annex.

| Assessment year |  | 2009 | 2010 |
| :---: | :---: | :---: | :---: |
| Assessment model |  | ICA | ICA |
| Tuning fleets | UK(E\&W)OTB | Series omitted | Series omitted |
|  | UK(E\&W)BTS Sept | 1989-2008 | 1989-2009 |
|  |  | ages 2-7 | ages 2-7 |
|  | UK(E\&W)BTS <br> March | Survey omitted | Survey omitted |
|  | UK(E\&W)BT | Series omitted | Series omitted |
|  | IR-OTB | Series omitted | Series omitted |
|  | UK(NI) GFS Mar | 1992-2008 | 1992-2009 |
|  |  | Biomass index | Biomass index |
|  | UK(NI) GFS Oct | 1992-2008 | 1992-2009 |
|  |  | Biomass index | Biomass index |
| Time-series weights |  | Full time-series - unweighted | Full time-series - unweighted |
| Num yrs for separable |  | 8 | 9 |
| Reference age |  | 5 | 5 |
| Terminal S |  | 1 | 1 |
| Catchability model fitted |  | linear | linear |
| SRR fitted |  | No | No |
| Landings number-at-age, range: |  | 2-9+ | 2-9+ |

Diagnostic output from FLICA is printed in Table 6.7.3.1 with estimates of fishing mortality and population numbers-at-age in Table 6.7.3.2 and 6.7.3.3 respectively. Patterns in the ICA residuals for UK (E\&W) beam trawl survey (linearly increasing ages 2 to 6) and UK (NI) GFS biomass survey (negative residuals between 2005 and 2008) have been noticed in recent years are present again in the assessment (Figure 6.7.3.1). Similarly, consistently positive residuals in the indices are apparent for age 5 throughout the separable period (2001-2009), while negative residuals are evident for ages 7 and 8. In the catch residuals, positive values are apparent in all years (except 2001) for age 5 data, while age 8 residuals are all negative. These patterns will need to be investigated at the benchmark meeting.

A retrospective analysis, with a constant separable period of nine years, was carried out and the results are shown in Figure 6.7.3.2. It can be seen that the assessment has a consistently biased retrospective pattern for SSB, recruitment and Fbar. Nevertheless, a general trend of increasing SSB and decreasing fishing mortality is evident.

## Comparison with previous assessments

Comparisons between this years and last year's ICA assessment are shown in Figure 6.7.3.4. The two assessments perform similarly in terms of temporal trends in SSB, recruitment and $\mathrm{F}_{\mathrm{b}}$. However, there is a slight difference in SSB and $\mathrm{F}_{\text {bar }}$ estimates between 2002 and 2006 in which the 2010WG estimates suggest a lower SSB $\vDash 6 \%$ di f ference) and a higher $\mathrm{F}_{\text {bar }}(\approx 9 \%)$. Nevertheless, the difference in estimates of $\mathrm{F}_{\text {bar }}$ in 2008 (+0.003) between the 2009WG and 2010WG assessments is very small. The SSB estimate for 2008 has been revised up $9 \%$ by 733 kt .

## State of the stock

Trends in Fbar, SSB, recruitment and landings, for the full time-series, are shown in Tables 6.7.3.4 and Figure 6.7.3.3. The update assessment estimates that fishing mortality rose to very high levels in the mid 1970s ( $\mathrm{F}_{\mathrm{bar}}>0.8$ ) but has declined from these levels over the subsequent 40 years. Indeed, since the early 1990s fishing mortality has shown a marked and almost continuous decline and in 2009 is estimated to be at the lowest level in the time-series (0.046). Spawning biomass levels show an oscillatory pattern over the time-series. High SSB levels occurred at the beginning of the time-series, and although it is estimated to have been steadily rising since 2000 it is still short of the earlier highs. Estimated recruitment levels have been variable over the time-series, but the levels declined markedly in the early 1990s and displayed only minor variations until 2008, which has the highest value since 1988. However, this has been followed by a recruitment estimate for 2009 of 3223 thousand fish, which is the lowest in the time-series and approximately half the previous minimum of 6005 thousand in 1990.

SSB in 2009 was above $\mathrm{B}_{\mathrm{pa}}$, and fishing mortality has been declining since the early 1990s and has been below $\mathrm{F}_{\mathrm{pa}}$ since 1998.

### 6.7.4 Short-term projections

A forecast is presented in this report as part of the usual update procedure. However, due to the consistent retrospective bias in the assessment, the Working Group considers any short-term forecast to be unreliable for this stock.

Population numbers for short-term forecasts were taken from the ICA output of survivors at ages 4 and above in 2010. Numbers-at-age 2 were taken as GM(90-07) ( 8.4 million). Because of the considerable uncertainty of the estimate of recruitment-at-age 2 in 2009, populations numbers-at-age 3 in 2010 have been overwritten with the $\mathrm{GM}(90-07)$ estimate depreciated for $\mathrm{F}_{\mathrm{sq}}$ and M (7.1 million at age 3 s in 2010).

The short-term forecast was run as status quo projection. Input data are shown in Table 6.7.4.1. The single option predicted forecast is given in Table 6.7.4.2, and the management option output is shown in Table 6.7.4.3 and summarised below.

| YEAR | LANDINGS (T) | SOURCE | SSB (T) JAN 1 ST | SOURCE |
| ---: | :---: | :--- | :--- | :--- |
| 2009 | 456 | WG Estimate | 7872 | ICA |
| 2010 | 779 | SQ Forecast | 10427 | SQ Forecast |
| 2011 | 839 | SQ Forecast | 11523 | SQ Forecast |

Proportions that the 2005 to 2009 year-classes will contribute to landings and SSB in 2010 and 2011 are shown in Table 6.7.4.4. Approximately $14 \%$ of the predicted land-
ings in 2010 and $33 \%$ of the predicted landings in 2011 rely on year classes for which geometric mean recruitment has been assumed.

The predicted landings for 2010 assuming status quo $F$ is $779 t$, and SSB is predicted to increase to 10427 t . The TAC for 2010 is 1627 t .

## Estimating recruiting year-class abundance

The update ICA estimates the strength of the 2007 year class at 3.2 million two year olds in 2009, which is below the geometric mean (1964-2007) of 11.7 million and the arithmetic mean (1964-2007) of 12.6 million. Considering the consistently low recruitment levels since the 1990s (Figure 6.7.3.3), GM90-07 (8.4 million) is used for the recruitment estimates in the short-term forecast.

Previous analyses have shown that recruitment estimates can be highly variable and dependant on model settings; therefore, recruitment is considered to be poorly estimated.

The recruitment estimates from various sources are shown below. Those used for the short-term forecasts are shown in bold.

| update assessment | ICA estimate | GM 90-07 |
| :--- | :---: | :---: |
| 2009 recruitment (000's)-at-age 2 | 3223 | 8355 |
| 2010 recruitment (000's)-at-age 2 |  | 8355 |
| 2011 recruitment (000's)-at-age 2 | 8355 |  |
| 2012 recruitment (000's)-at-age 2 | 8355 |  |

### 6.7.5 Medium-term projections

There are no medium-term projections for this stock.

### 6.7.6 MSY explorations

## Modelling approach

MSY reference points were explored using the Cefas ADMB module presented at WKFRAME 2010 and based on sen and sum files created from stock assessment outputs. The model applied assumes a single species harvest scenario with no densitydependent variation in growth or mortality rates at high stock abundance. The models used do not include uncertainty due to ecosystem effects and multi-species interactions affecting growth, maturity and natural mortality. Therefore the variability estimated at low fishing mortality rates is likely to be underestimated and the potential yields over estimated.

Stock and recruitment curves, assuming a smooth hockey stick or the traditional Ricker or Beverton-Holt models, were fitted to the data and the diagnostic output evaluated to determine the appropriate function for the estimation of $\mathrm{F}_{\text {msy }}$ or its proxies. Delta AICc values were also computed to guide model selection. Variability in the model, determined through MCMC re-sampling, reflects variance in the stockrecruit relationship and in the biological parameters of growth and maturity. Stocks of plaice do not generally show a strong stock-recruit relationship therefore, a priori, we favour a smooth hockey stick approach. For comparison, conventional YPR analyses were also conducted.

## Results

From 1000 projections with each model, 525 runs were acceptable (in which the bounds of the fit were not violated.) based on the Ricker model, 225 on the BevertonHolt (BH) model and 542 with the smooth hockey stick (HS) (Figure 6.7.3.6). Therefore, for this stock, the BH model appears particularly problematic and the deterministic fit is outside the confidence intervals. The fiftieth percentile estimate of $\mathrm{F}_{\text {msy }}$ is greatest when assuming a HS model $(0.45$, CV $59 \%)$ and lowest when based on the BH model ( 0.20 , CV $43 \%$ ), while the Ricker model estimate ( 0.35 , CV $23 \%$ ) is midway between and associated with a favourable coefficient of variability (Table 7.6.4.6, Figures 6.7.3.6-8). Notably, when assuming a HS model, $\mathrm{F}_{\text {msy }}$ is equal to the current $\mathrm{F}_{\mathrm{pa}}$ estimate.

## Conclusions

Given the noisy data, large rejection rate for all S-R models and relatively high CVs the stock-recruit relationship of plaice is not well captured by any of the models and the underlying data do not support the provision of absolute estimates of $\mathrm{F}_{\mathrm{msy}}$.

### 6.7.7 Biological reference points

## Precautionary approach reference points

Biological reference points were proposed for this stock by the 1998 Working Group as below:

| Flim | No proposal |  |
| :--- | :--- | :--- |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.45 | (on the basis of Fmed and long- <br> term considerations) |
| $\mathrm{B}_{\lim }$ | No proposal |  |
| $\mathrm{B}_{\mathrm{pa}}$ | 3100 t | (on the basis of Bloss and <br> evidence of highrecruitments at <br> low SSBs) |

## Yield-per-recruit analysis

Yield-per-recruit analyses were performed and presented in Table 6.7.4.5 and Figure 6.7.3.5, but given the uncertainties associated with the short-term forecast of this stock, the results should be treated with caution. $\mathrm{F}_{\max }$ was calculated as 0.625 , and $\mathrm{F}_{0.1}$ as 0.146 . Notably, $\mathrm{F}_{\max }$ is poorly defined and $\mathrm{F}_{0.1}$ forgoes a great amount of yield. The yield-per-recruit analysis done as part of the MSY evaluations illustrates significant uncertainty (Figure 6.7.3.8) and both $\mathrm{F}_{\max }$ and $\mathrm{F}_{0.1}$ are poorly defined (CV of 58 and $92 \%$ respectively, Table 6.7.4.6), highlighting that the use of a deterministic YPR is inappropriate for this stock.

### 6.7.8 Management plans

There are no management plans for this stock.

### 6.7.9 Uncertainties and bias in assessment and forecast

It has been noted in previous years that aspects of this assessment appear to be deteriorating. Specific concerns in recent years have been the contradictory signals pro-
vided by the surveys, a retrospective bias in estimates of $\mathrm{Fbar}_{\text {and }}$ and SSB and the lack of discard data and contrast in the strength of incoming year classes.

Discard levels in this fishery are estimated to be very high and fish at the younger ages may be subject to substantially higher mortality levels than currently estimated. The landings of young fish represent only a small proportion of those caught and the lack of adequate information on mortality rates at these ages seriously impairs the ability to estimate recruitment levels in the population. There are no sufficiently reliable estimates of discard levels for the entire time-series of landings for this stock, to enable inclusion in the assessment.

The only age based tuning data in this assessment is restricted to the area where the increase in the plaice stock appears to be most dramatic. Further work needs to be carried out to determine to which degree the rise in SSB predicted by the UK (E\&W) beam trawl survey is representative of the stock as a whole.

Landings weights-at-age calculations for this stock were problematic this year and last year for ages greater than 12 due to the low number of sampled fish. Also UK (E\&W) and Irish values of weight-at-age show very different values in these older fish. There is evidence of a decline in weight-at-age from the raw commercial landings data and survey data.

### 6.7.10 Recommendations for next benchmark

2010 ICES Review: The 2010 ICES Review Group raised concerns regarding the stock definition, the paucity of ecosystem information and lack of discard information in the analysis. The RG noted that multiple fisheries catch plaice below the legal size and discards are greater than retained catch levels in all fisheries, which may be the root cause of the retrospective patterns seen in the assessment. The RG agreed with the WG that the forecasts are not reliable. Indeed, the RG concluded further that forecasts should not be included in the assessment or ICES advice. These RG suggested that these issues should be addressed as part of the benchmark process.


### 6.7.11 Management considerations

The high level of discarding (typically up to $80 \%$ in number) in this fishery indicates a mismatch between the minimum landing size and the mesh size of the gear being used. Any measures that effect a reduction in discards will result in increased future yield. However, decreasing the mesh size may not have the desired result since the market demand for plaice is poor and small plaice are particularly undesirable.

Status quo F (average 2006-2009) is estimated to be 0.0701; below $\mathrm{F}_{0.1}$ and well below Fpa. SSB in 2009 is estimated at 7872 t , and at 10427 t in 2010, both of which are well above $\mathbf{B}_{\mathrm{pa}}(3100 \mathrm{t})$. However, given the poor fit of the assessment model, estimates of fishing mortality and stock biomass should be interpreted with caution.

Whilst the precise levels of Fbar and SSB are considered poorly estimated, the overall state of the stock is consistently estimated to have low fishing mortality ( $<\mathrm{F}_{\mathrm{pa}}$ ) and high spawning biomass ( $>\mathrm{Bp}_{\mathrm{pa}}$ ). Therefore the stock is considered to be within safe biological limits.

A fishing mortality of $\mathrm{F}_{\mathrm{pa}}(0.45)$ forecasts that landings in 2011 would be 4480 tonnes (Table 6.7.4.3). This however requires a substantial increase in $\mathrm{Fbar}_{\mathrm{b}}(\mathrm{F}$ multiplier $=$ 6.42), and the landings would be far greater than the current TAC level, which is currently not met by the fishery. However, due to the consistent retrospective bias in the
assessment the Working Group considers any short-term forecast to be unreliable and the results should therefore be treated with caution.

Table 6.7.2.1. Nominal landings of plaice in Division VIIa as officially reported to ICES.


Table 6.7.2.2. Irish Sea plaice: English standardised lpue and effort, Belgian beam trawl lpue and effort and Irish otter trawl lpue and effort-series.

| Year | CPUE |  | LPUE |  |  |  |  | Effort ('O00hrs) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\text { UK }(E \& W)^{4}$ |  | English ${ }^{1}$ |  | $\begin{aligned} & \frac{\text { Belgian }^{3}}{\text { Beam }} \\ & \text { Trawl } \\ & \hline \end{aligned}$ | $\text { Irish }^{7}$ |  | $\text { English }^{2}$ |  | $\begin{aligned} & \frac{\text { Belgian }^{5}}{\text { Beam }} \\ & \text { Trawl } \\ & \hline \end{aligned}$ | Irish |  |
|  | Beam tr | awl survey | Otter | Beam |  | Otter |  | Otter | Beam |  | Otter |  |
|  | March | September | Trawl | Trawl |  | Trawl | Trawl | Trawl | Trawl |  | Trawl | Trawl |
| 1972 |  |  | 6.96 |  | 9.8 |  |  | 128.4 |  | 6.8 |  |  |
| 1973 |  |  | 6.33 |  | 9.0 |  |  | 147.6 |  | 16.5 |  |  |
| 1974 |  |  | 7.45 |  | 10.4 |  |  | 115.2 |  | 14.2 |  |  |
| 1975 |  |  | 7.71 |  | 10.7 |  |  | 130.7 |  | 16.2 |  |  |
| 1976 |  |  | 5.03 |  | 5.8 |  |  | 122.3 |  | 15.1 |  |  |
| 1977 |  |  | 4.82 |  | 5.3 |  |  | 101.9 |  | 13.4 |  |  |
| 1978 |  |  | 6.77 | 4.88 | 6.9 |  |  | 89.1 | 0.9 | 12.0 |  |  |
| 1979 |  |  | 7.18 | 15.23 | 8.0 |  |  | 89.9 | 1.7 | 13.7 |  |  |
| 1980 |  |  | 8.24 | 8.98 | 8.6 |  |  | 107.0 | 4.3 | 20.8 |  |  |
| 1981 |  |  | 6.87 | 4.91 | 7.1 |  |  | 107.1 | 6.4 | 26.7 |  |  |
| 1982 |  |  | 4.92 | 1.77 | 4.4 |  |  | 127.2 | 5.5 | 21.3 |  |  |
| 1983 |  |  | 5.32 | 3.08 | 7.8 |  |  | 88.1 | 2.8 | 18.5 |  |  |
| 1984 |  |  | 7.77 | 6.98 | 6.8 |  |  | 103.1 | 4.1 | 13.6 |  |  |
| 1985 |  |  | 9.97 | 25.70 | 8.8 |  |  | 102.9 | 7.4 | 21.9 |  |  |
| 1986 |  |  | 9.27 | 4.21 | 8.7 |  |  | 90.3 | 17.0 | 38.3 |  |  |
| 1987 |  |  | 7.20 | 3.57 | 8.2 |  |  | 130.6 | 22.0 | 43.2 |  |  |
| 1988 |  | 392 | 5.02 | 3.05 | 6.3 |  |  | 132.0 | 18.6 | 32.7 |  |  |
| 1989 |  | 253 | 5.51 | 13.59 | 6.2 |  |  | 139.5 | 25.3 | 36.7 |  |  |
| 1990 |  | 239 | 5.93 | 12.02 | 7.2 |  |  | 117.1 | 31.0 | 38.3 |  |  |
| 1991 |  | 157 | 4.79 | 10.56 | 7.5 |  |  | 107.3 | 25.8 | 15.4 |  |  |
| 1992 |  | 188 | 4.20 | 9.99 | 11.9 |  |  | 96.8 | 23.4 | 23.0 |  |  |
| 1993 | 91 | 235 | 3.97 | 9.50 | 5.0 |  |  | 78.9 | 21.5 | 24.4 |  |  |
| 1994 | 128 | 225 | 4.90 | 7.79 | 9.2 |  |  | 43.0 | 20.1 | 31.6 |  |  |
| 1995 | 134 | 169 | 5.08 | 7.69 | 9.5 | 3.2 | 17.0 | 43.1 | 20.9 | 27.1 | 80.3 | 8.6 |
| 1996 | - ${ }^{6}$ | 210 | 5.37 | 12.96 | 11.8 | 4.1 | 18.9 | 42.2 | 13.3 | 22.2 | 64.8 | 6.3 |
| 1997 | 147 | 262 | 5.25 | 7.66 | 13.9 | 3.1 | 13.7 | 39.9 | 10.8 | 29.3 | 92.2 | 9.0 |
| 1998 | 113 | 249 | 5.00 | 5.66 | 12.3 | 3.7 | 22.2 | 36.9 | 10.4 | 23.8 | 93.5 | 11.6 |
| 1999 | - ${ }^{6}$ | 264 | 5.38 | 7.76 | 12.0 | 2.3 | 23.2 | 22.9 | 11.0 | 22.1 | 110.3 | 14.7 |
| 2000 | .$^{6}$ | 357 | 5.02 | 13.04 | 11.6 | 2.0 | 13.8 | 27.0 | 6.3 | 18.2 | 82.7 | 11.4 |
| 2001 |  | 281 | 3.35 | 8.33 | 13.6 | 2.5 | 10.8 | 33.0 | 12.5 | 28.5 | 77.5 | 13.1 |
| 2002 |  | 340 | 5.66 | 5.46 | 10.7 | 2.8 | 7.9 | 24.8 | 8.0 | 36.2 | 77.9 | 17.7 |
| 2003 |  | 503 | 2.60 | 3.76 | 8.8 | 4.1 | 9.5 | 23.9 | 14.0 | 23.0 | 73.8 | 18.7 |
| 2004 |  | 540 | 3.17 | 4.20 | 14.9 | 2.1 | 8.6 | 23.5 | 7.4 | 27.6 | 72.5 | 14.2 |
| 2005 |  | 367 | 4.85 | 4.67 | 15.3 | 2.0 | 8.0 | 16.7 | 11.6 | 31.8 | 68.3 | 14.7 |
| 2006 |  | 356 | 6.50 | 2.19 | 11.6 | 1.4 | 6.3 | 5.2 | 4.6 | 28.1 | 64.9 | 11.9 |
| 2007 |  | 432 | 17.94 | 4.22 | 7.3 | 1.2 | 6.1 | 4.4 | 3.2 | 22.3 | 73.2 | 14.0 |
| 2008 |  | 416 | 9.03 | 4.47 | 10.6 | 0.9 | 5.2 | 2.7 | 1.3 | 11.1 | 58.8 | 9.5 |
| 2009 |  | 467 | 6.49 | 1.21 | 16.7 | 1.0 | 3.8 | 1.5 | 0.46 | 8.9 | 41.5 | 7.6 |

L Whole weight ( kg ) per corrected hour fished, weighted by area
2 Corrected for fishing power (GRT)
$3 \mathrm{Kg} / \mathrm{hr}$
$+\mathrm{Kg} / 100 \mathrm{~km}$
; Corrected for fishing power (HP) [uncorrected data for 2007-2008, replaced at 2010WG].
; Carhelmar survey, $\mathrm{Kg} / 100 \mathrm{~km}$ not available
7 All years updated in 2007 due to slight historical differences
تishing power corrections are detailed in Appendix 2 of the 2000 working group report

Table 6.7.2.3. Irish Sea plaice: UK (NI) index of relative SSB trends by region.

| JK(NI) GFS Mar | Estimated m | an abun | nce | Estimated | dard e |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Combined | West | East | Combined | West | East |
| Year | Str1-7 | Str1-3 | Str4-7 | Str1-7 | Str1-3 | Str4-7 |
| 1992 | 9.59 | 6.40 | 10.54 | 4.39 | 2.13 | 5.66 |
| 1993 | 13.27 | 21.40 | 10.85 | 2.22 | 5.56 | 2.36 |
| 1994 | 10.09 | 5.38 | 11.50 | 2.56 | 1.83 | 3.27 |
| 1995 | 7.59 | 6.56 | 7.89 | 1.39 | 1.66 | 1.74 |
| 1996 | 7.96 | 14.41 | 6.04 | 1.68 | 5.94 | 1.28 |
| 1997 | 13.73 | 15.80 | 13.11 | 3.99 | 6.78 | 4.76 |
| 1998 | 12.50 | 19.61 | 10.38 | 3.62 | 10.88 | 3.39 |
| 1999 | 9.37 | 19.10 | 6.46 | 2.34 | 7.42 | 2.09 |
| 2000 | 15.79 | 35.36 | 9.96 | 5.40 | 22.56 | 1.97 |
| 2001 | 13.52 | 23.78 | 10.46 | 2.11 | 6.21 | 2.02 |
| 2002 | 13.36 | 25.65 | 9.70 | 3.24 | 8.93 | 3.25 |
| 2003 | 26.79 | 55.52 | 18.23 | 8.36 | 32.38 | 4.95 |
| 2004 | 10.55 | 8.60 | 11.13 | 4.77 | 5.23 | 7.58 |
| 2005 | 15.86 | 27.20 | 12.48 | 3.54 | 8.59 | 3.82 |
| 2006 | 9.57 | 16.33 | 7.55 | 1.80 | 6.15 | 1.45 |
| 2007 | 8.73 | 21.76 | 4.84 | 1.81 | 7.00 | 1.06 |
| 2008 | 6.33 | 9.26 | 5.46 | 0.90 | 5.71 | 1.01 |
| 2009 | 11.00 | 17.85 | 8.96 | 1.89 | 4.61 | 2.03 |
| 2010 | 22.67 | 16.49 | 24.51 | 3.80 | 4.49 | 4.75 |
| JK(NI) GFS Oct | Estimated m | an abun | nce | Estimated | ndard er |  |
|  | Combined | West | East | Combined | West | East |
| Year | Str1-7 | Str1-3 | Str4-7 | Str1-7 | Str1-3 | Str4-7 |
| 1991 | 0.81 | 3.38 | 0.04 | 0.39 | 1.71 | 0.03 |
| 1992 | 4.83 | 2.76 | 5.45 | 0.85 | 1.26 | 1.04 |
| 1993 | 4.64 | 2.91 | 5.16 | 0.95 | 1.18 | 1.18 |
| 1994 | 9.20 | 8.65 | 9.36 | 2.27 | 3.74 | 2.72 |
| 1995 | 4.77 | 8.31 | 3.72 | 1.28 | 3.52 | 1.29 |
| 1996 | 8.69 | 9.95 | 8.32 | 2.15 | 5.67 | 2.22 |
| 1997 | 8.22 | 7.67 | 8.38 | 2.18 | 2.80 | 2.71 |
| 1998 | 5.39 | 4.21 | 5.74 | 1.45 | 2.39 | 1.75 |
| 1999 | 6.90 | 4.91 | 7.50 | 2.29 | 3.12 | 2.82 |
| 2000 | 10.50 | 2.84 | 12.78 | 6.42 | 1.16 | 8.33 |
| 2001 | 13.93 | 4.03 | 16.88 | 6.45 | 1.96 | 8.35 |
| 2002 | 9.98 | 6.63 | 10.98 | 3.80 | 3.45 | 4.82 |
| 2003 | 18.65 | 10.09 | 21.20 | 5.41 | 4.87 | 6.87 |
| 2004 | 8.49 | 2.52 | 10.28 | 1.90 | 1.10 | 2.44 |
| 2005 | 11.58 | 3.88 | 13.88 | 4.39 | 2.39 | 5.66 |
| 2006 | 7.20 | 2.59 | 8.57 | 1.98 | 1.47 | 2.53 |
| 2007 | 8.48 | 6.09 | 9.19 | 1.69 | 2.55 | 2.05 |
| 2008 | 11.28 | 4.66 | 13.26 | 3.06 | 2.50 | 3.91 |
| 2009 | 14.83 | 5.36 | 17.66 | 3.25 | 3.71 | 4.07 |

Table 6.7.2.4. Irish Sea plaice: tuning fleet data available. Figures shown in bold are those used in the assessment.

Irish Sea plaice, 2010
101
UK BT SURVEY (Sept-Trad) - Prime stations only
19892009
110.750 .85

18
$\begin{array}{lllllllll}129.710 & 309 & 441 & 530 & 77 & 13 & 44 & 3 & 0\end{array}$
$\begin{array}{lllllllll}128.969 & 1688 & 405 & 176 & 90 & 54 & 30 & 3 & 1\end{array}$
$\begin{array}{lllllllll}123.780 & 591 & 481 & 68 & 47 & 4 & 4 & 24 & 3\end{array}$
$\begin{array}{lllllllll}129.525 & 1043 & 470 & 267 & 23 & 19 & 14 & 14 & 3\end{array}$
$\begin{array}{lllllllll}131.192 & 1106 & 812 & 136 & 101 & 16 & 8 & 21 & 4\end{array}$
$\begin{array}{lllllllll}124.892 & 815 & 608 & 307 & 68 & 33 & 12 & 17 & 8\end{array}$
$\begin{array}{lllllllll}126.004 & 1283 & 387 & 179 & 84 & 16 & 18 & 0 & 1\end{array}$
$\begin{array}{lllllllll}126.004 & 1701 & 601 & 124 & 74 & 49 & 9 & 11 & 1\end{array}$
$\begin{array}{lllllllll}126.004 & 1363 & 668 & 322 & 65 & 50 & 23 & 8 & 7\end{array}$
$\begin{array}{lllllllll}126.004 & 1167 & 767 & 212 & 95 & 34 & 23 & 14 & 3\end{array}$
$\begin{array}{lllllllll}126.004 & 1189 & 965 & 344 & 113 & 38 & 17 & 7 & 7\end{array}$
$\begin{array}{lllllllll}126.004 & 2112 & 659 & 298 & 141 & 73 & 22 & 7 & 3\end{array}$
$\begin{array}{lllllllll}126.004 & 1468 & 663 & 218 & 130 & 89 & 28 & 10 & 7\end{array}$
$126.00417341615 \quad 647 \quad 243 \quad 79 \quad 51 \quad 1617$
$126.00414801842 \quad 827 \quad 296122 \quad 62 \quad 3910$
$126.004181611871184404261 \quad 57 \quad 5714$
$122.298 \quad 8691295 \quad 6664992971111717$
$126.0041120 \quad 840 \quad 722411178 \quad 835916$
$126.004 \quad 26671255 \quad 525417196 \quad 954537$
$122.29812931893 \quad 628 \quad 339 \quad 243 \quad 76 \quad 5533$
126.0041460108312253101892516531

UK(E+W)TRAWL FLEET (calculated using ABBT age compositions)

19872009

1101

114


UK (E+W)BEAM TRAWL FLEET

19872009
1101

114
$21.9970 .0 \quad 1.1 \quad 27.1113 .136 .031 .3 \quad 2.9 \quad 6.7 \quad 1.93 .10 .60 .10 .20 .1$ $18.564 \quad 0.0 \quad 2.0 \quad 48.0 \quad 23.7 \quad 24.4 \quad 13.2 \quad 8.5 \quad 1.4 \quad 2.6 \quad 1.61 .5 \quad 0.6 \quad 0.8 \quad 0.3$ $25.291 \quad 3.1132 .8 \quad 297.5163 .452 .642 .425 .1 \quad 16.1 \quad 4.35 .33 .35 .7 \quad 2.611 .1$ $31.0032 .2136 .2391 .9361 .178 .230 .217 .2 \quad 8.4 \quad 3.61 .51 .93 .81 .4 \quad 0.5$ $25.83817 .3282 .5182 .9174 .591 .8 \quad 35.911 .211 .8 \quad 3.54 .7$ 0.2 1.0 0.6 0.3 $23.3993 .9141 .5335 .6 \quad 79.6 \quad 64.6 \quad 45.5 \quad 18.6 \quad 8.0 \quad 12.27 .14 .0 \quad 0.20 .71 .0$
$21.5030 .6 \quad 73.4112 .8 \quad 95.2 \quad 23.3 \quad 24.232 .011 .8 \quad 4.5 \quad 7.12 .21 .20 .0 \quad 0.4$ $20.14513 .4151 .8186 .1 \quad 39.926 .0 \quad 6.8 \quad 6.6 \quad 7.8 \quad 3.51 .20 .91 .20 .20 .0$ $20.9325 .2183 .4229 .1100 .633 .116 .1 \quad 3.9 \quad 1.7 \quad 3.31 .0 \quad 0.9 \quad 0.5 \quad 0.1 \quad 0.2$
 $10.760 \quad 0.9 \quad 98.6 \quad 69.5 \quad 39.0 \quad 30.2 \quad 13.5 \quad 3.7 \quad 3.2 \quad 0.5 \quad 0.4 \quad 0.3 \quad 0.2 \quad 0.1 \quad 0.1$ $10.3860 .3 \quad 63.5103 .7 \quad 32.612 .0 \quad 9.7 \quad 6.3 \quad 2.7 \quad 1.8 \quad 0.30 .20 .5 \quad 0.20 .0$ $11.0164 .851 .3124 .4 \quad 80.424 .412 .510 .5 \quad 5.6 \quad 0.9 \quad 0.8 \quad 0.2 \quad 0.2 \quad 0.20 .1$ $6.2750 .0 \quad 25.2 \quad 61.4 \quad 46.6 \quad 27.9 \quad 7.3 \quad 6.5 \quad 4.5 \quad 1.9 \quad 0.7 \quad 0.7 \quad 0.7 \quad 0.1 \quad 0.1$ $12.495 \quad 1.5 \quad 20.6 \quad 47.5 \quad 56.6 \quad 42.7 \quad 20.8 \quad 7.0 \quad 4.5 \quad 2.51 .20 .4 \quad 0.1 \quad 0.10 .0$ $8.017 \quad 0.0 \quad 11.5 \quad 33.1 \quad 21.0 \quad 18.8 \quad 14.9 \quad 8.0 \quad 2.3 \quad 1.31 .4 \quad 0.4 \quad 0.4 \quad 0.0 \quad 0.0$ $13.9960 .0 \quad 11.4 \quad 45.5 \quad 47.7 \quad 20.9 \quad 10.0 \quad 8.7 \quad 5.4 \quad 1.7 \quad 0.30 .0 \quad 0.3 \quad 0.0 \quad 0.1$ $7.396 \quad 0.2 \quad 18.0 \quad 29.4 \quad 11.711 .9 \quad 5.1 \quad 1.7 \quad 1.4 \quad 1.0 \quad 0.3 \quad 0.20 .10 .0 \quad 0.0$ $11.406 \quad 0.1 \quad 6.5 \quad 11.0 \quad 24.020 .7 \quad 9.2 \quad 3.4 \quad 1.6 \quad 1.30 .4 \quad 0.4 \quad 0.10 .10 .0$ $4.649 \quad 0.2 \quad 2.7 \quad 8.1 \quad 4.9 \quad 8.2 \quad 3.8 \quad 2.6 \quad 0.9 \quad 0.6 \quad 0.5 \quad 0.2 \quad 0.2 \quad 0.1 \quad 0.0$ $3.197 \quad 0.0 \quad 0.2 \quad 3.2 \quad 7.2 \quad 4.5 \quad 5.3 \quad 1.8 \quad 1.3 \quad 0.30 .30 .1 \quad 0.10 .0 \quad 0.0$ $1.300 \quad 0.0 \quad 0.0 \quad 1.4 \quad 3.5 \quad 3.9 \quad 2.1 \quad 1.7 \quad 0.8 \quad 0.30 .1 \quad 0.1 \quad 0.0 \quad 0.0 \quad 0.0$ $0.4620 .00 .0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0$ UK BT SURVEY (March) - Prime stations only

19931999
110.150 .25

18
$\begin{array}{lllllllll}126.931 & 480 & 662 & 141 & 71 & 12 & 8 & 11 & 3\end{array}$
$\begin{array}{lllllllll}115.442 & 361 & 662 & 370 & 98 & 47 & 5 & 7 & 10\end{array}$
$\begin{array}{lllllllll}126.189 & 859 & 647 & 340 & 120 & 29 & 28 & 0 & 10\end{array}$
$\begin{array}{lllllllll}134.343 & 1559 & 908 & 295 & 98 & 49 & 16 & 8 & 1\end{array}$
$\begin{array}{lllllllll}121.742 & 967 & 905 & 351 & 63 & 39 & 31 & 10 & 13\end{array}$
$\begin{array}{lllllllll}130.081 & 648 & 957 & 217 & 82 & 24 & 23 & 12 & 1\end{array}$
$\begin{array}{lllllllll}130.822 & 570 & 770 & 389 & 98 & 26 & 11 & 9 & 6\end{array}$
IR-JPS : Irish Juvenile Plaice Survey 2nd Qtr - Effort min. towed - Plaice No. at age

19912004
110.370 .43

17

```
555 185 206 60 21 
570
600
585
570
675
675
675
660 0 0 0 0 0 0 0 0 0
645
675
```




```
660
IR-OTB : Irish Otter trawl - Effort in hours - VIIa Plaice numbers at age -
Year
1 9 9 5 2 0 0 9
1101
212
70682
58166
75029
81073
93221
\begin{tabular}{llllllllllll}
64320 & 11 & 92 & 98 & 88 & 24 & 10 & 8 & 3 & 1 & 4 & 0
\end{tabular}
\begin{tabular}{llllllllllll}
77541 & 55 & 90 & 97 & 104 & 100 & 38 & 16 & 11 & 3 & 1 & 0
\end{tabular}
77863
```



```
72507
68336
64876
73157
58812 4 4 16 
```

| 41469 | 2 | 17 | 22 | 20 | 23 | 10 | 9 | 2 | 2 | 1 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

UK(NI) GFS Spring and autumn spawning biomass indices

2162
'Year''VPA' 'DARDS' 'DARDA'
199219.594 .83
1993113.274 .64
1994110.099 .20
199517.594 .77
199617.968 .69
1997113.738 .22
1998112.505 .39
199919.376 .90
2000115.7910 .50
2001113.5213 .93
2002113.369 .98
2003126.7918 .65
2004110.558 .49
2005115.8611 .58
200619.577 .20
200718.738 .48
200816.3311 .28
2009111.0014 .83
2010122.67

Table 6.7.2.5. Irish Sea plaice: Landings numbers-at-ages 1 to 15+ (thousands).

|  | Ages |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1964 | 0 | 997 | 1911 | 1680 | 446 | 851 | 480 | 140 | 26 | 155 | 30 | 2 | 1 | 1 | 10 |
| 1965 | 28 | 1416 | 3155 | 2841 | 1115 | 555 | 309 | 300 | 17 | 20 | 5 | 2 | 1 | 1 | 1 |
| 1966 | 0 | 120 | 4303 | 3605 | 2182 | 620 | 588 | 386 | 181 | 13 | 20 | 7 | 7 | 3 | 6 |
| 1967 | 0 | 164 | 1477 | 5593 | 4217 | 995 | 642 | 267 | 210 | 176 | 86 | 35 | 5 | 6 | 1 |
| 1968 | 0 | 171 | 1961 | 3410 | 4641 | 1611 | 319 | 113 | 135 | 24 | 17 | 3 | 4 | 1 | 1 |
| 1969 | 59 | 430 | 2317 | 2932 | 2080 | 2227 | 779 | 184 | 58 | 100 | 80 | 22 | 9 | 4 | 1 |
| 1970 | 9 | 803 | 2278 | 2179 | 1877 | 1028 | 899 | 239 | 64 | 29 | 52 | 51 | 20 | 3 | 2 |
| 1971 | 0 | 427 | 3392 | 3882 | 1683 | 1371 | 491 | 497 | 244 | 60 | 65 | 36 | 11 | 9 | 1 |
| 1972 | 0 | 142 | 3254 | 5136 | 1461 | 752 | 555 | 627 | 353 | 169 | 55 | 40 | 38 | 19 | 12 |
| 1973 | 0 | 925 | 4091 | 5233 | 2682 | 642 | 345 | 238 | 183 | 238 | 129 | 40 | 14 | 11 | 17 |
| 1974 | 7 | 1200 | 2530 | 2694 | 2125 | 1045 | 191 | 139 | 56 | 47 | 95 | 40 | 5 | 5 | 5 |
| 1975 | 18 | 1370 | 4313 | 1902 | 1158 | 933 | 152 | 119 | 81 | 94 | 47 | 72 | 18 | 16 | 4 |
| 1976 | 23 | 2553 | 4333 | 2425 | 902 | 563 | 391 | 198 | 59 | 79 | 47 | 22 | 58 | 11 | 5 |
| 1977 | 565 | 4124 | 2767 | 2470 | 839 | 236 | 150 | 112 | 63 | 21 | 15 | 8 | 8 | 10 | 3 |
| 1978 | 22 | 3063 | 5169 | 1535 | 542 | 202 | 98 | 54 | 52 | 43 | 10 | 9 | 4 | 4 | 2 |
| 1979 | 12 | 3380 | 5679 | 1835 | 363 | 187 | 109 | 61 | 68 | 68 | 17 | 5 | 6 | 4 | 6 |
| 1980 | 3 | 2783 | 6738 | 2560 | 646 | 312 | 125 | 64 | 24 | 54 | 16 | 13 | 7 | 5 | 5 |
| 1981 | 22 | 1742 | 5939 | 2984 | 837 | 222 | 105 | 53 | 52 | 41 | 28 | 35 | 13 | 3 | 11 |
| 1982 | 27 | 715 | 3288 | 3082 | 1358 | 330 | 137 | 69 | 44 | 36 | 11 | 15 | 11 | 14 | 13 |
| 1983 | 51 | 2924 | 2494 | 3211 | 1521 | 648 | 211 | 110 | 53 | 30 | 13 | 15 | 9 | 11 | 11 |
| 1984 | 41 | 3159 | 5179 | 1182 | 1054 | 459 | 299 | 113 | 60 | 13 | 22 | 15 | 10 | 6 | 13 |
| 1985 | 4 | 2357 | 6152 | 3301 | 614 | 429 | 262 | 181 | 78 | 36 | 21 | 8 | 7 | 3 | 6 |
| 1986 | 31 | 1652 | 5280 | 2942 | 1287 | 344 | 371 | 112 | 92 | 54 | 24 | 9 | 5 | 3 | 9 |
| 1987 | 62 | 3717 | 5317 | 5252 | 1341 | 1072 | 123 | 121 | 75 | 74 | 25 | 8 | 10 | 12 | 13 |
| 1988 | 46 | 2923 | 5040 | 2552 | 1400 | 750 | 316 | 84 | 112 | 44 | 41 | 28 | 38 | 21 | 37 |
| 1989 | 24 | 1735 | 5945 | 2671 | 854 | 436 | 214 | 153 | 56 | 47 | 26 | 38 | 18 | 7 | 19 |
| 1990 | 15 | 1019 | 2715 | 2935 | 1132 | 465 | 259 | 98 | 51 | 22 | 15 | 15 | 9 | 6 | 7 |
| 1991 | 180 | 2008 | 1506 | 1929 | 1205 | 465 | 182 | 122 | 49 | 34 | 5 | 6 | 3 | 3 | 4 |
| 1992 | 151 | 1958 | 3209 | 1435 | 1358 | 903 | 388 | 118 | 74 | 44 | 27 | 15 | 9 | 3 | 4 |
| 1993 | 28 | 910 | 1649 | 1357 | 474 | 556 | 377 | 179 | 42 | 50 | 16 | 8 | 2 | 3 | 2 |
| 1994 | 98 | 1146 | 2173 | 1309 | 644 | 318 | 245 | 134 | 86 | 18 | 6 | 9 | 6 | 1 | 3 |
| 1995 | 21 | 961 | 1703 | 1936 | 764 | 318 | 138 | 70 | 47 | 23 | 9 | 4 | 1 | 1 | 3 |
| 1996 | 37 | 856 | 1345 | 1196 | 943 | 370 | 128 | 44 | 25 | 37 | 14 | 7 | 5 | 1 | 2 |
| 1997 | 28 | 830 | 1590 | 1513 | 1003 | 482 | 285 | 139 | 42 | 53 | 12 | 7 | 1 | 2 | 1 |
| 1998 | 5 | 691 | 1739 | 1025 | 612 | 476 | 403 | 177 | 91 | 52 | 25 | 17 | 19 | 2 | 1 |
| 1999 | 68 | 803 | 1505 | 1294 | 696 | 280 | 196 | 117 | 69 | 43 | 6 | 4 | 1 | 0 | 1 |
| 2000 | 0 | 450 | 1174 | 1284 | 685 | 212 | 219 | 102 | 55 | 19 | 14 | 7 | 2 | 2 | 2 |
| 2001 | 14 | 374 | 1138 | 1083 | 767 | 409 | 178 | 90 | 45 | 18 | 6 | 2 | 4 | 0 | 0 |
| 2002 | 1 | 206 | 940 | 1482 | 842 | 539 | 318 | 96 | 48 | 17 | 4 | 3 | 0 | 0 | 0 |
| 2003 | 0 | 286 | 1031 | 1314 | 707 | 415 | 253 | 127 | 48 | 22 | 12 | 7 | 1 | 3 | 0 |
| 2004 | 7 | 198 | 967 | 1104 | 705 | 246 | 114 | 88 | 74 | 11 | 11 | 1 | 1 | 0 | 0 |
| 2005 | 6 | 228 | 708 | 1177 | 890 | 461 | 204 | 92 | 55 | 37 | 12 | 12 | 4 | 2 | 1 |
| 2006 | 5 | 180 | 620 | 550 | 684 | 346 | 220 | 87 | 53 | 46 | 20 | 6 | 2 | 1 | 1 |
| 2007 | 0 | 64 | 350 | 859 | 506 | 401 | 150 | 114 | 27 | 14 | 5 | 3 | 0 | 0 | 0 |
| 2008 | 1 | 99 | 386 | 389 | 409 | 215 | 141 | 61 | 36 | 9 | 7 | 3 | 1 | 1 | 0 |
| 2009 | 0 | 13 | 204 | 374 | 351 | 272 | 116 | 73 | 26 | 12 | 4 | 2 | 1 | 1 | 1 |

Table 6.7.2.6. Irish Sea plaice: Landings weights-at-ages 1 to $15+(\mathbf{k g})$.

|  | Ages |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1964 | 0.000 | 0.190 | 0.292 | 0.413 | 0.463 | 0.597 | 0.831 | 1.042 | 1.155 | 0.552 | 1.358 | 1.015 | 1.544 | 1.605 | 1.654 |
| 1965 | 0.070 | 0.177 | 0.269 | 0.388 | 0.556 | 0.653 | 0.690 | 0.719 | 0.801 | 1.198 | 1.167 | 0.971 | 1.477 | 1.535 | 1.581 |
| 1966 | 0.000 | 0.152 | 0.223 | 0.316 | 0.418 | 0.532 | 0.697 | 0.691 | 0.939 | 0.983 | 1.074 | 1.071 | 1.233 | 1.281 | 1.320 |
| 1967 | 0.000 | 0.133 | 0.218 | 0.299 | 0.382 | 0.516 | 0.518 | 0.759 | 0.791 | 0.682 | 0.783 | 0.514 | 1.152 | 1.198 | 1.234 |
| 1968 | 0.000 | 0.149 | 0.213 | 0.313 | 0.413 | 0.509 | 0.584 | 0.777 | 0.893 | 0.957 | 1.017 | 0.887 | 1.174 | 1.220 | 1.257 |
| 1969 | 0.056 | 0.146 | 0.215 | 0.311 | 0.405 | 0.541 | 0.643 | 0.787 | 0.897 | 0.744 | 0.723 | 1.097 | 1.185 | 1.231 | 1.269 |
| 1970 | 0.058 | 0.149 | 0.219 | 0.324 | 0.417 | 0.523 | 0.648 | 0.685 | 0.908 | 0.925 | 0.877 | 0.603 | 1.231 | 1.279 | 1.318 |
| 1971 | 0.000 | 0.140 | 0.207 | 0.295 | 0.396 | 0.489 | 0.595 | 0.753 | 0.654 | 0.852 | 0.731 | 1.079 | 1.153 | 1.198 | 1.235 |
| 1972 | 0.000 | 0.143 | 0.235 | 0.332 | 0.432 | 0.560 | 0.737 | 0.712 | 0.959 | 1.071 | 1.144 | 1.208 | 1.288 | 1.339 | 1.379 |
| 1973 | 0.000 | 0.143 | 0.218 | 0.316 | 0.415 | 0.491 | 0.645 | 0.694 | 0.791 | 0.898 | 0.927 | 0.863 | 1.204 | 1.252 | 1.290 |
| 1974 | 0.063 | 0.158 | 0.246 | 0.334 | 0.445 | 0.514 | 0.686 | 0.847 | 0.964 | 1.052 | 1.108 | 1.048 | 1.326 | 1.378 | 1.420 |
| 1975 | 0.072 | 0.185 | 0.275 | 0.398 | 0.531 | 0.644 | 0.749 | 0.924 | 1.147 | 1.169 | 1.359 | 1.360 | 1.533 | 1.593 | 1.641 |
| 1976 | 0.060 | 0.150 | 0.228 | 0.323 | 0.419 | 0.525 | 0.590 | 0.719 | 0.797 | 0.842 | 0.834 | 1.003 | 1.267 | 1.317 | 1.357 |
| 1977 | 0.059 | 0.153 | 0.226 | 0.340 | 0.430 | 0.510 | 0.592 | 0.738 | 0.840 | 1.016 | 0.945 | 1.100 | 1.252 | 1.301 | 1.340 |
| 1978 | 0.071 | 0.185 | 0.268 | 0.391 | 0.525 | 0.672 | 0.720 | 0.910 | 1.035 | 1.049 | 1.264 | 1.329 | 1.497 | 1.556 | 1.603 |
| 1979 | 0.069 | 0.176 | 0.262 | 0.376 | 0.557 | 0.668 | 0.794 | 0.915 | 0.997 | 0.968 | 1.274 | 1.227 | 1.471 | 1.529 | 1.575 |
| 1980 | 0.066 | 0.177 | 0.255 | 0.365 | 0.483 | 0.517 | 0.671 | 0.884 | 1.047 | 1.072 | 1.259 | 1.273 | 1.403 | 1.458 | 1.503 |
| 1981 | 0.069 | 0.176 | 0.267 | 0.376 | 0.512 | 0.592 | 0.678 | 0.863 | 1.097 | 0.804 | 1.276 | 1.310 | 1.309 | 1.509 | 1.554 |
| 1982 | 0.201 | 0.274 | 0.284 | 0.348 | 0.421 | 0.545 | 0.650 | 0.651 | 0.780 | 0.777 | 1.185 | 1.164 | 1.147 | 1.164 | 1.744 |
| 1983 | 0.232 | 0.261 | 0.290 | 0.319 | 0.368 | 0.426 | 0.484 | 0.552 | 0.629 | 0.716 | 0.803 | 0.910 | 1.026 | 1.161 | 1.316 |
| 1984 | 0.260 | 0.290 | 0.330 | 0.380 | 0.470 | 0.560 | 0.660 | 0.760 | 0.870 | 0.980 | 1.100 | 1.240 | 1.420 | 1.630 | 1.940 |
| 1985 | 0.290 | 0.310 | 0.340 | 0.390 | 0.470 | 0.540 | 0.630 | 0.730 | 0.840 | 0.940 | 1.060 | 1.200 | 1.380 | 1.600 | 1.900 |
| 1986 | 0.270 | 0.280 | 0.340 | 0.420 | 0.500 | 0.540 | 0.630 | 0.830 | 0.920 | 1.020 | 1.210 | 1.480 | 1.420 | 1.720 | 1.610 |
| 1987 | 0.260 | 0.290 | 0.315 | 0.370 | 0.440 | 0.520 | 0.610 | 0.720 | 0.820 | 0.950 | 1.080 | 1.210 | 1.360 | 1.520 | 1.700 |
| 1988 | 0.230 | 0.260 | 0.300 | 0.370 | 0.460 | 0.550 | 0.680 | 0.820 | 0.960 | 1.120 | 1.300 | 1.480 | 1.690 | 1.900 | 2.130 |
| 1989 | 0.227 | 0.272 | 0.321 | 0.374 | 0.430 | 0.491 | 0.555 | 0.623 | 0.694 | 0.770 | 0.849 | 0.932 | 1.019 | 1.109 | 1.205 |
| 1990 | 0.200 | 0.257 | 0.316 | 0.376 | 0.439 | 0.504 | 0.570 | 0.639 | 0.709 | 0.781 | 0.856 | 0.932 | 1.010 | 1.091 | 1.173 |
| 1991 | 0.247 | 0.267 | 0.295 | 0.332 | 0.377 | 0.431 | 0.494 | 0.566 | 0.646 | 0.735 | 0.832 | 0.938 | 1.053 | 1.176 | 1.309 |
| 1992 | 0.169 | 0.218 | 0.274 | 0.337 | 0.407 | 0.484 | 0.568 | 0.658 | 0.756 | 0.860 | 0.971 | 1.089 | 1.213 | 1.345 | 1.483 |
| 1993 | 0.260 | 0.270 | 0.292 | 0.328 | 0.375 | 0.436 | 0.508 | 0.594 | 0.691 | 0.802 | 0.925 | 1.060 | 1.208 | 1.368 | 1.541 |
| 1994 | 0.156 | 0.207 | 0.268 | 0.338 | 0.416 | 0.504 | 0.600 | 0.706 | 0.821 | 0.945 | 1.077 | 1.219 | 1.370 | 1.530 | 1.698 |
| 1995 | 0.201 | 0.229 | 0.266 | 0.312 | 0.366 | 0.429 | 0.501 | 0.581 | 0.670 | 0.768 | 0.874 | 0.990 | 1.114 | 1.246 | 1.387 |
| 1996 | 0.144 | 0.203 | 0.268 | 0.338 | 0.414 | 0.496 | 0.584 | 0.677 | 0.776 | 0.881 | 0.992 | 1.108 | 1.230 | 1.358 | 1.492 |
| 1997 | 0.134 | 0.184 | 0.239 | 0.299 | 0.362 | 0.430 | 0.502 | 0.579 | 0.660 | 0.745 | 0.834 | 0.928 | 1.027 | 1.129 | 1.236 |
| 1998 | 0.202 | 0.222 | 0.252 | 0.294 | 0.346 | 0.410 | 0.484 | 0.569 | 0.665 | 0.773 | 0.891 | 1.020 | 1.160 | 1.310 | 1.472 |
| 1999 | 0.174 | 0.213 | 0.257 | 0.309 | 0.366 | 0.430 | 0.501 | 0.577 | 0.661 | 0.751 | 0.847 | 0.949 | 1.058 | 1.174 | 1.296 |
| 2000 | 0.000 | 0.222 | 0.257 | 0.302 | 0.357 | 0.422 | 0.497 | 0.581 | 0.676 | 0.780 | 0.894 | 1.018 | 1.152 | 1.296 | 1.450 |
| 2001 | 0.142 | 0.205 | 0.269 | 0.337 | 0.407 | 0.479 | 0.554 | 0.632 | 0.712 | 0.795 | 0.880 | 0.968 | 1.058 | 1.151 | 1.247 |
| 2002 | 0.185 | 0.225 | 0.271 | 0.324 | 0.383 | 0.449 | 0.521 | 0.600 | 0.685 | 0.776 | 0.874 | 0.978 | 1.089 | 1.206 | 1.329 |
| 2003 | 0.000 | 0.244 | 0.289 | 0.340 | 0.395 | 0.455 | 0.520 | 0.590 | 0.665 | 0.745 | 0.830 | 0.920 | 1.014 | 1.114 | 1.219 |
| 2004 | 0.207 | 0.230 | 0.261 | 0.300 | 0.348 | 0.404 | 0.468 | 0.542 | 0.623 | 0.713 | 0.811 | 0.918 | 1.033 | 1.157 | 1.289 |
| 2005 | 0.172 | 0.212 | 0.254 | 0.299 | 0.345 | 0.394 | 0.445 | 0.499 | 0.554 | 0.612 | 0.672 | 0.734 | 0.799 | 0.865 | 0.934 |
| 2006 | 0.227 | 0.232 | 0.249 | 0.278 | 0.320 | 0.374 | 0.440 | 0.518 | 0.609 | 0.712 | 0.827 | 0.954 | 1.094 | 1.246 | 1.410 |
| 2007 | 0.000 | 0.215 | 0.247 | 0.283 | 0.325 | 0.371 | 0.422 | 0.479 | 0.540 | 0.606 | 0.677 | 0.753 | 0.834 | 0.920 | 1.011 |
| 2008 | 0.000 | 0.224 | 0.233 | 0.252 | 0.280 | 0.318 | 0.365 | 0.421 | 0.486 | 0.560 | 0.644 | 0.737 | 0.840 | 0.951 | 1.072 |
| 2009 | 0.000 | 0.174 | 0.224 | 0.272 | 0.315 | 0.355 | 0.391 | 0.424 | 0.453 | 0.478 | 0.499 | 0.517 | 0.531 | 0.542 | 0.549 |

Table 6.7.2.7. Irish Sea plaice: Stock weights-at-ages 1 to $15+(\mathbf{k g})$.

|  | Ages |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| 1964 | 0.02 | 0.11 | 0.23 | 0.35 | 0.41 | 0.55 | 0.77 | 0.98 | 1.09 | 0.54 | 1.31 | 0.99 | 1.51 | 1.54 | 1.63 |
| 1965 | 0.02 | 0.11 | 0.21 | 0.33 | 0.48 | 0.59 | 0.64 | 0.68 | 0.77 | 1.15 | 1.13 | 0.95 | 1.44 | 1.48 | 1.56 |
| 1966 | 0.02 | 0.09 | 0.18 | 0.27 | 0.37 | 0.48 | 0.64 | 0.65 | 0.88 | 0.95 | 1.04 | 1.04 | 1.20 | 1.23 | 1.30 |
| 1967 | 0.02 | 0.08 | 0.17 | 0.25 | 0.34 | 0.46 | 0.48 | 0.72 | 0.75 | 0.66 | 0.76 | 0.51 | 1.13 | 1.15 | 1.22 |
| 1968 | 0.02 | 0.08 | 0.17 | 0.26 | 0.36 | 0.46 | 0.54 | 0.73 | 0.84 | 0.92 | 0.98 | 0.86 | 1.15 | 1.17 | 1.24 |
| 1969 | 0.02 | 0.08 | 0.17 | 0.26 | 0.36 | 0.49 | 0.59 | 0.74 | 0.84 | 0.72 | 0.70 | 1.06 | 1.16 | 1.19 | 1.25 |
| 1970 | 0.02 | 0.09 | 0.18 | 0.27 | 0.37 | 0.47 | 0.60 | 0.65 | 0.85 | 0.89 | 0.85 | 0.59 | 1.20 | 1.23 | 1.30 |
| 1971 | 0.02 | 0.08 | 0.16 | 0.25 | 0.35 | 0.44 | 0.55 | 0.71 | 0.63 | 0.82 | 0.71 | 1.04 | 1.13 | 1.15 | 1.22 |
| 1972 | 0.02 | 0.09 | 0.19 | 0.28 | 0.38 | 0.50 | 0.68 | 0.67 | 0.90 | 1.03 | 1.10 | 1.17 | 1.26 | 1.29 | 1.36 |
| 1973 | 0.02 | 0.09 | 0.17 | 0.27 | 0.36 | 0.45 | 0.60 | 0.66 | 0.75 | 0.87 | 0.90 | 0.84 | 1.18 | 1.20 | 1.27 |
| 1974 | 0.02 | 0.09 | 0.19 | 0.28 | 0.39 | 0.47 | 0.63 | 0.80 | 0.91 | 1.01 | 1.07 | 1.02 | 1.30 | 1.33 | 1.40 |
| 1975 | 0.02 | 0.11 | 0.22 | 0.34 | 0.46 | 0.58 | 0.70 | 0.87 | 1.08 | 1.13 | 1.31 | 1.32 | 1.50 | 1.53 | 1.62 |
| 1976 | 0.02 | 0.09 | 0.18 | 0.27 | 0.37 | 0.48 | 0.55 | 0.68 | 0.76 | 0.81 | 0.81 | 0.97 | 1.24 | 1.27 | 1.34 |
| 1977 | 0.02 | 0.09 | 0.18 | 0.29 | 0.38 | 0.46 | 0.55 | 0.70 | 0.79 | 0.98 | 0.91 | 1.07 | 1.22 | 1.25 | 1.32 |
| 1978 | 0.02 | 0.11 | 0.21 | 0.33 | 0.46 | 0.60 | 0.67 | 0.86 | 0.98 | 1.01 | 1.22 | 1.29 | 1.46 | 1.50 | 1.58 |
| 1979 | 0.02 | 0.10 | 0.21 | 0.32 | 0.48 | 0.60 | 0.73 | 0.86 | 0.94 | 0.94 | 1.23 | 1.19 | 1.44 | 1.47 | 1.55 |
| 1980 | 0.02 | 0.10 | 0.20 | 0.31 | 0.42 | 0.47 | 0.62 | 0.83 | 0.98 | 1.03 | 1.22 | 1.23 | 1.37 | 1.40 | 1.48 |
| 1981 | 0.02 | 0.10 | 0.21 | 0.32 | 0.45 | 0.54 | 0.63 | 0.81 | 1.03 | 0.78 | 1.23 | 1.27 | 1.28 | 1.45 | 1.53 |
| 1982 | 0.02 | 0.09 | 0.21 | 0.31 | 0.41 | 0.48 | 0.57 | 0.66 | 0.75 | 0.85 | 0.95 | 1.05 | 1.15 | 1.26 | 1.37 |
| 1983 | 0.02 | 0.09 | 0.21 | 0.30 | 0.35 | 0.40 | 0.46 | 0.52 | 0.59 | 0.68 | 0.77 | 0.86 | 0.97 | 1.09 | 1.24 |
| 1984 | 0.02 | 0.10 | 0.23 | 0.35 | 0.43 | 0.52 | 0.61 | 0.71 | 0.82 | 0.93 | 1.04 | 1.17 | 1.33 | 1.53 | 1.79 |
| 1985 | 0.02 | 0.10 | 0.24 | 0.36 | 0.43 | 0.51 | 0.59 | 0.68 | 0.79 | 0.89 | 1.00 | 1.13 | 1.29 | 1.49 | 1.75 |
| 1986 | 0.02 | 0.12 | 0.26 | 0.38 | 0.44 | 0.52 | 0.61 | 0.72 | 0.83 | 0.96 | 1.12 | 1.26 | 1.41 | 1.56 | 1.72 |
| 1987 | 0.02 | 0.10 | 0.24 | 0.35 | 0.41 | 0.48 | 0.56 | 0.66 | 0.77 | 0.89 | 1.01 | 1.15 | 1.29 | 1.44 | 1.61 |
| 1988 | 0.25 | 0.26 | 0.29 | 0.34 | 0.40 | 0.48 | 0.59 | 0.70 | 0.84 | 1.00 | 1.17 | 1.36 | 1.57 | 1.79 | 2.03 |
| 1989 | 0.21 | 0.25 | 0.30 | 0.35 | 0.40 | 0.46 | 0.52 | 0.59 | 0.66 | 0.73 | 0.81 | 0.89 | 0.98 | 1.06 | 1.16 |
| 1990 | 0.17 | 0.23 | 0.29 | 0.35 | 0.41 | 0.47 | 0.54 | 0.60 | 0.67 | 0.75 | 0.82 | 0.89 | 0.97 | 1.05 | 1.13 |
| 1991 | 0.24 | 0.26 | 0.28 | 0.31 | 0.35 | 0.40 | 0.46 | 0.53 | 0.61 | 0.69 | 0.78 | 0.88 | 0.99 | 1.11 | 1.24 |
| 1992 | 0.15 | 0.19 | 0.25 | 0.31 | 0.37 | 0.45 | 0.53 | 0.61 | 0.71 | 0.81 | 0.91 | 1.03 | 1.15 | 1.28 | 1.41 |
| 1993 | 0.26 | 0.26 | 0.28 | 0.31 | 0.35 | 0.40 | 0.47 | 0.55 | 0.64 | 0.75 | 0.86 | 0.99 | 1.13 | 1.29 | 1.45 |
| 1994 | 0.13 | 0.18 | 0.24 | 0.30 | 0.38 | 0.46 | 0.55 | 0.65 | 0.76 | 0.88 | 1.01 | 1.15 | 1.29 | 1.45 | 1.61 |
| 1995 | 0.19 | 0.21 | 0.25 | 0.29 | 0.34 | 0.40 | 0.46 | 0.54 | 0.63 | 0.72 | 0.82 | 0.93 | 1.05 | 1.18 | 1.32 |
| 1996 | 0.12 | 0.17 | 0.23 | 0.30 | 0.38 | 0.45 | 0.54 | 0.63 | 0.73 | 0.83 | 0.94 | 1.05 | 1.17 | 1.29 | 1.42 |
| 1997 | 0.11 | 0.16 | 0.21 | 0.27 | 0.33 | 0.40 | 0.47 | 0.54 | 0.62 | 0.70 | 0.79 | 0.88 | 0.98 | 1.08 | 1.18 |
| 1998 | 0.20 | 0.21 | 0.24 | 0.27 | 0.32 | 0.38 | 0.45 | 0.53 | 0.62 | 0.72 | 0.83 | 0.95 | 1.09 | 1.23 | 1.39 |
| 1999 | 0.16 | 0.19 | 0.23 | 0.28 | 0.34 | 0.40 | 0.47 | 0.54 | 0.62 | 0.71 | 0.80 | 0.90 | 1.00 | 1.12 | 1.23 |
| 2000 | 0.00 | 0.21 | 0.24 | 0.28 | 0.33 | 0.39 | 0.46 | 0.54 | 0.63 | 0.73 | 0.84 | 0.96 | 1.08 | 1.22 | 1.37 |
| 2001 | 0.11 | 0.17 | 0.24 | 0.30 | 0.37 | 0.44 | 0.52 | 0.59 | 0.67 | 0.75 | 0.84 | 0.92 | 1.01 | 1.11 | 1.20 |
| 2002 | 0.17 | 0.20 | 0.25 | 0.30 | 0.35 | 0.42 | 0.48 | 0.56 | 0.64 | 0.73 | 0.82 | 0.93 | 1.03 | 1.15 | 1.27 |
| 2003 | 0.00 | 0.22 | 0.27 | 0.31 | 0.37 | 0.42 | 0.49 | 0.55 | 0.63 | 0.70 | 0.79 | 0.87 | 0.97 | 1.06 | 1.17 |
| 2004 | 0.20 | 0.22 | 0.24 | 0.28 | 0.32 | 0.38 | 0.44 | 0.50 | 0.58 | 0.67 | 0.76 | 0.86 | 0.98 | 1.09 | 1.22 |
| 2005 | 0.15 | 0.19 | 0.23 | 0.28 | 0.32 | 0.37 | 0.42 | 0.47 | 0.53 | 0.58 | 0.64 | 0.70 | 0.77 | 0.83 | 0.90 |
| 2006 | 0.23 | 0.23 | 0.24 | 0.26 | 0.30 | 0.35 | 0.41 | 0.48 | 0.56 | 0.66 | 0.77 | 0.89 | 1.02 | 1.17 | 1.33 |
| 2007 | 0.00 | 0.20 | 0.23 | 0.27 | 0.30 | 0.35 | 0.40 | 0.45 | 0.51 | 0.57 | 0.64 | 0.71 | 0.79 | 0.88 | 0.97 |
| 2008 | 0.22 | 0.23 | 0.24 | 0.27 | 0.30 | 0.34 | 0.39 | 0.45 | 0.52 | 0.60 | 0.69 | 0.79 | 0.89 | 1.01 | 1.14 |
| 2009 | 0.00 | 0.15 | 0.20 | 0.25 | 0.29 | 0.34 | 0.37 | 0.41 | 0.44 | 0.47 | 0.49 | 0.51 | 0.53 | 0.54 | 0.55 |

Table 6.7.3.1. Irish Sea plaice: Final ICA diagnostics and output.

FLICA CONFIGURATION SETTINGS

```
sep.2 : NA
sep.gradual : TRUE
sr : FALSE
sr.age : 2
lambda.age : 1 1 1 1 1 1 1 0
lambda.yr : 1 1 11111111
lambda.sr : 0
index.model : linear linear linear
index.cor : 1 -925596313493178307362200 -925596313493178307362200
sep.nyr : 9
sep.age : 5
sep.sel : 1
```

FLR, R SOFTWARE VERSIONS
R version 2.8.1 (2008-12-22)
Package : FLICA
Version : 1.4-12
Packaged : 2009-10-08 15:16:26 UTC; mpa
Built : R 2.9.1; ; 2009-10-08 15:16:27 UTC; windows
Package : FLAssess
Version : 1.99-102
Packaged : Mon Mar 23 08:18:19 2009; mpa
Built : R 2.8.0; i386-pc-mingw32; 2009-03-23 08:18:21; windows
Package : FLCore
Version : 2.2
Packaged : Tue May 19 19:23:18 2009; Administrator
Built : R 2.8.1; i386-pc-mingw32; 2009-05-19 19:23:22; windows

FITTED SELECTION PATTERN

Units : NA

| Year |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 |
| 3 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 |
| 4 | 1.078 | 1.078 | 1.078 | 1.078 | 1.078 | 1.078 | 1.078 | 1.078 | 1.078 |
| 5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 6 | 1.116 | 1.116 | 1.116 | 1.116 | 1.116 | 1.116 | 1.116 | 1.116 | 1.116 |
| 7 | 0.944 | 0.944 | 0.944 | 0.944 | 0.944 | 0.944 | 0.944 | 0.944 | 0.944 |
| 8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 6.7.3.1. (cont).

## FIT PARAMETERS

|  | Value | Std.dev | Lower.95.pct.CL | Upper.95.pct.CL |
| :---: | :---: | :---: | :---: | :---: |
| F2001 | 0.360523 | 0.1676 | 0.259598 | 0.500685 |
| F2002 | 0.357308 | 0.1671 | 0.257501 | 0.495801 |
| F2003 | 0.324909 | 0.1714 | 0.232218 | 0.454599 |
| F2004 | 0.209357 | 0.1746 | 0.148687 | 0.294784 |
| F2005 | 0.228515 | 0.1728 | 0.162852 | 0.320653 |
| F2006 | 0.153648 | 0.1714 | 0.109809 | 0.214987 |
| F2007 | 0.105777 | 0.169 | 0.07595 | 0.147317 |
| F2008 | 0.070282 | 0.1684 | 0.050529 | 0.097758 |
| F2009 | 0.049362 | 0.1751 | 0.035024 | 0.069569 |
| Selectivity at age 2 | 0.104984 | 0.1653 | 0.075935 | 0.145145 |
| Selectivity at age 3 | 0.536738 | 0.1506 | 0.399567 | 0.721001 |
| Selectivity at age 4 | 1.07779 | 0.1381 | 0.82224 | 1.412764 |
| Selectivity at age 6 | 1.115684 | 0.1245 | 0.874134 | 1.423981 |
| Selectivity at age 7 | 0.944497 | 0.1233 | 0.74179 | 1.202597 |
| Terminal year popage 2 | 3221.501647 | 0.3434 | 1643.480701 | 6314.691045 |
| Terminal year popage 3 | 10269.86787 | 0.2415 | 6396.687947 | 16488.24938 |
| Terminal year popage 4 | 6548.676293 | 0.1989 | 4434.18381 | 9671.489281 |
| Terminal year popage 5 | 5311.593314 | 0.1716 | 3794.359981 | 7435.515784 |
| Terminal year popage 6 | 4715.670137 | 0.1553 | 3478.249168 | 6393.31565 |
| Terminal year popage 7 | 2602.214258 | 0.1527 | 1929.049085 | 3510.288616 |
| Terminal year popage 8 | 2174.319325 | 0.1521 | 1613.686586 | 2929.729086 |
| Last true age pop2001 | 324.443093 | 0.3326 | 169.059059 | 622.642297 |
| Last true age pop2002 | 378.80891 | 0.2484 | 232.799521 | 616.393838 |
| Last true age pop2003 | 596.277139 | 0.2181 | 388.83798 | 914.381941 |
| Last true age pop2004 | 702.691156 | 0.2119 | 463.862361 | 1064.485723 |
| Last true age pop2005 | 687.45457 | 0.1959 | 468.293967 | 1009.181879 |
| Last true age pop2006 | 869.850702 | 0.184 | 606.491832 | 1247.568729 |
| Last true age pop2007 | 1386.144405 | 0.167 | 999.197834 | 1922.938827 |
| Last true age pop2008 | 1518.268688 | 0.1602 | 1109.202669 | 2078.195331 |
| Index 2 biomass | 0.002294 | 0.0794 | 0.001963 | 0.00268 |
| Index 3 biomass | 0.00175 | 0.0794 | 0.001498 | 0.002045 |
| Index 1 age 2 numbers | 0.000917 | 0.166 | 0.000662 | 0.00127 |
| Index 1 age 3 numbers | 0.000529 | 0.1646 | 0.000383 | 0.000731 |
| Index 1 age 4 numbers | 0.000347 | 0.1644 | 0.000251 | 0.000478 |
| Index 1 age 5 numbers | 0.000251 | 0.165 | 0.000182 | 0.000347 |
| Index 1 age 6 numbers | 0.000238 | 0.1666 | 0.000172 | 0.00033 |
| Index 1 age 7 numbers | 0.000232 | 0.1741 | 0.000165 | 0.000327 |

Table 6.7.3.1. (cont).

INDEX RESIDUALS

UK(E\&W) BTS
Units: NA

| year |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 2 | -0.8963 | -0.41818 | -0.6 | -0.52202 | -0.205 | -0.21978 | -0.539 | -0.03933 |
| 3 | -0.0183 | -0.77834 | -1.14 | -0.00854 | -0.788 | -0.16396 | -0.4784 | -0.73395 |
| 4 | -0.6597 | -0.63775 | -1.02 | -1.07001 | 0.119 | -0.47548 | -0.4349 | -0.44153 |
| 5 | -1.056 | -0.00762 | -2.76 | -0.81417 | -0.456 | 0.00211 | -1.0505 | -0.04304 |
| 6 | 0.6982 | 0.62051 | -1.96 | -0.67469 | -0.76 | 0.12918 | 0.0212 | -1.07532 |
| 7 | -0.9646 | -1.27605 | 1.17 | -0.01745 | 0.62 | 0.99827 | -99 | 0.00381 |
|  | year |  |  |  |  |  |  |  |
| age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 2 | -0.176 | -0.0732 | 0.3037 | -0.1109 | -0.363 | 0.688 | 0.664 | 0.392 |
| 3 | 0.353 | -0.3698 | 0.0245 | 0.0126 | -0.379 | 0.443 | 0.835 | 0.989 |
| 4 | -0.299 | 0.1048 | -0.0753 | 0.0048 | 0.0851 | 0.62 | 0.523 | 0.878 |
| 5 | 0.155 | -0.0158 | 0.1517 | 0.2791 | 0.3432 | 0.392 | 0.709 | 1.076 |
| 6 | -0.178 | 0.1159 | -0.1513 | 0.1159 | -0.116 | 0.351 | 0.685 | 0.373 |
| 7 | -0.597 | 0.1433 | -0.4746 | -0.4291 | -0.197 | -0.179 | 0.554 | 0.977 |
| year |  |  |  |  |  |  |  |  |
| age | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |  |
| 2 | 0.24 | -0.0861 | 0.3218 | 0.4713 | 1.1678 |  |  |  |
| 3 | 0.616 | 0.366 | 0.1621 | 0.3613 | 0.6985 |  |  |  |
| 4 | 0.918 | 0.8041 | 0.4685 | 0.3702 | 0.2193 |  |  |  |
| 5 | 1.269 | 0.4703 | 0.6226 | 0.4785 | 0.2519 |  |  |  |
| 6 | 0.67 | 0.3208 | 0.1411 | -0.0374 | 0.7128 |  |  |  |
| 7 | -0.442 | 0.3214 | -0.0314 | -0.1531 | -0.025 |  |  |  |

NI SSB Spring
Units: NA

| year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| all | -0.532 | -0.417 | 0.291 | -0.329 | 0.215 | 0.238 | -0.23 | 0.0397 | 0.417 |
|  | year |  |  |  |  |  |  |  |  |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| all | 0.557 | 0.149 | 0.648 | -0.111 | 0.114 | -0.472 | -0.339 | -0.314 | 0.0735 |

NI SSB Autumn
Units: NA
year

| age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| all | 0.364 | 0.113 | -0.135 | -0.143 | 0.481 | 0.341 | 0.0753 | 0.555 | 0.257 |
|  | year |  |  |  |  |  |  |  |  |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| all | 0.171 | 0.74 | -0.165 | 0.158 | -0.457 | -0.581 | -1.16 | -0.496 | 0 |

## Table 6.7.3.1. (cont).

## CATCH RESIDUALS

Units: Thousands NA
year

age |  |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 2 | 0.132 | -0.2947 | -0.02554 | 0.2171 | 0.00096 | 0.3008 | -0.352 | 0.203 | -0.1885 |
|  | 0.1097 | -0.3373 | 0.00245 | 0.1925 | -0.03724 | -0.061 | -0.137 | 0.366 | -0.2158 |  |
|  | 4 | -0.1443 | 0.0884 | -0.2186 | 0.1319 | -0.09974 | -0.3268 | 0.159 | -0.133 | 0.1561 |
| 5 | -0.089 | 0.1821 | -0.00433 | 0.0776 | 0.25074 | 0.1518 | 0.295 | 0.113 | 0.375 |  |
| 6 | -0.1979 | -0.0443 | -0.0601 | -0.3277 | -0.19337 | -0.0851 | 0.138 | -0.049 | 0.1323 |  |
| 7 | 0.0803 | 0.2185 | -0.06266 | -0.3405 | -0.09071 | -0.0467 | -0.124 | -0.118 | 0.0369 |  |
| 8 | -0.0345 | -0.117 | -0.20977 | -0.3551 | -0.36731 | -0.2968 | -0.142 | -0.466 | -0.3024 |  |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 6.7.3.2. Irish Sea plaice: Final ICA population numbers-at-age (thousands).

| age | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 21672 | 29108 | 15094 | 13596 | 10907 | 12508 | 18548 | 17283 | 11825 | 8746 | 11710 | 11514 |
| 3 | 11209 | 18283 | 24485 | 13274 | 11904 | 9513 | 10689 | 15695 | 14927 | 10354 | 6887 | 9258 |
| 4 | 5390 | 8147 | 13252 | 17674 | 10385 | 8716 | 6263 | 7342 | 10736 | 10184 | 5353 | 3739 |
| 5 | 2779 | 3206 | 4564 | 8372 | 10433 | 6014 | 4982 | 3513 | 2887 | 4721 | 4145 | 2231 |
| 6 | 2587 | 2046 | 1798 | 2008 | 3486 | 4912 | 3385 | 2661 | 1543 | 1196 | 1686 | 1692 |
| 7 | 1677 | 1496 | 1294 | 1014 | 851 | 1586 | 2274 | 2039 | 1080 | 666 | 461 | 521 |
| 8 | 516 | 1037 | 1037 | 597 | 302 | 456 | 678 | 1175 | 1347 | 439 | 268 | 230 |
| 9 | 829 | 163 | 637 | 1161 | 494 | 679 | 627 | 1007 | 1474 | 1167 | 488 | 642 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 2 | 9664 | 15014 | 16168 | 20089 | 18194 | 13554 | 7389 | 18737 | 18777 | 19800 | 14385 | 17600 |
| 3 | 8924 | 6176 | 9448 | 11464 | 14642 | 13522 | 10384 | 5881 | 13871 | 13686 | 15345 | 11205 |
| 4 | 4179 | 3866 | 2890 | 3556 | 4861 | 6686 | 6437 | 6128 | 2882 | 7452 | 6385 | 8663 |
| 5 | 1539 | 1445 | 1129 | 1130 | 1440 | 1921 | 3139 | 2829 | 2436 | 1450 | 3522 | 2912 |
| 6 | 897 | 524 | 499 | 495 | 662 | 673 | 921 | 1514 | 1089 | 1175 | 712 | 1918 |
| 7 | 630 | 272 | 244 | 253 | 264 | 296 | 389 | 508 | 737 | 536 | 640 | 310 |
| 8 | 320 | 194 | 101 | 125 | 123 | 117 | 164 | 216 | 253 | 373 | 231 | 222 |
| 9 | 454 | 222 | 232 | 356 | 238 | 404 | 342 | 279 | 311 | 328 | 404 | 398 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 2 | 18827 | 11496 | 6605 | 10234 | 9041 | 9897 | 8267 | 7147 | 6691 | 8302 | 8453 | 7474 |
| 3 | 12120 | 13952 | 8566 | 4901 | 7191 | 6181 | 7922 | 6255 | 5436 | 5130 | 6583 | 6847 |
| 4 | 4968 | 6033 | 6812 | 5052 | 2935 | 3377 | 3935 | 4988 | 3950 | 3559 | 3059 | 4207 |
| 5 | 2789 | 2023 | 2853 | 3296 | 2675 | 1262 | 1725 | 2263 | 2611 | 2382 | 1741 | 1753 |
| 6 | 1329 | 1166 | 995 | 1470 | 1794 | 1104 | 675 | 927 | 1291 | 1433 | 1174 | 971 |
| 7 | 701 | 479 | 626 | 448 | 868 | 748 | 460 | 302 | 524 | 798 | 819 | 596 |
| 8 | 160 | 326 | 225 | 312 | 227 | 407 | 311 | 179 | 139 | 345 | 441 | 350 |
| 9 | 610 | 450 | 287 | 266 | 338 | 280 | 300 | 225 | 287 | 292 | 516 | 371 |
| year |  |  |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| 2 | 7399 | 9360 | 7970 | 9281 | 7777 | 10195 | 8835 | 8743 | 11666 | 3223 |  |  |
| 3 | 5874 | 6139 | 7994 | 6808 | 7956 | 6748 | 8827 | 7711 | 7669 | 10271 |  |  |
| 4 | 4660 | 4107 | 4487 | 5852 | 5072 | 6306 | 5294 | 7209 | 6461 | 6550 |  |  |
| 5 | 2518 | 2929 | 2470 | 2708 | 3657 | 3590 | 4372 | 3979 | 5705 | 5313 |  |  |
| 6 | 903 | 1591 | 1811 | 1533 | 1735 | 2631 | 2533 | 3325 | 3175 | 4717 |  |  |
| 7 | 598 | 602 | 944 | 1078 | 946 | 1218 | 1808 | 1893 | 2621 | 2603 |  |  |
| 8 | 345 | 325 | 380 | 597 | 704 | 688 | 871 | 1387 | 1519 | 2175 |  |  |
| 9 | 341 | 262 | 253 | 355 | 549 | 637 | 960 | 517 | 891 | 1035 |  |  |

        year
    age 2010
2 NA
32843
48871
55508
64485
73959
82204
92710

Table 6.7.3.3. Irish Sea plaice: Final ICA fishing mortality-at-age.

| age | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.05 | 0.053 | 0.00847 | 0.0129 | 0.0168 | 0.0372 | 0.047 | 0.0266 | 0.0128 | 0.119 | 0.115 |  |
| 3 | 0.199 | 0.202 | 0.20594 | 0.1255 | 0.1917 | 0.298 | 0.256 | 0.2597 | 0.2623 | 0.54 | 0.491 |  |
| 4 | 0.4 | 0.459 | 0.33925 | 0.4072 | 0.4262 | 0.4392 | 0.458 | 0.8134 | 0.7015 | 0.779 | 0.755 |  |
| 5 | 0.186 | 0.458 | 0.70095 | 0.7561 | 0.6332 | 0.4547 | 0.507 | 0.7029 | 0.7615 | 0.91 | 0.776 |  |
| 6 | 0.427 | 0.338 | 0.45288 | 0.7382 | 0.6678 | 0.6501 | 0.387 | 0.7819 | 0.7206 | 0.833 | 1.053 |  |
| 7 | 0.36 | 0.247 | 0.65249 | 1.0927 | 0.5036 | 0.729 | 0.54 | 0.2941 | 0.7792 | 0.789 | 0.575 |  |
| 8 | 0.339 | 0.365 | 0.49923 | 0.6372 | 0.5035 | 0.5544 | 0.465 | 0.5908 | 0.6744 | 0.844 | 0.789 |  |
| 9 | $\begin{aligned} & 0.339 \\ & \text { year } \end{aligned}$ | 0.365 | 0.49923 | 0.6372 | 0.5035 | 0.5544 | 0.465 | 0.5908 | 0.6744 | 0.844 | 0.789 |  |
| age | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 2 | 0.135 | 0.328 | 0.343 | 0.224 | 0.196 | 0.177 | 0.146 | 0.108 | 0.181 | 0.196 | 0.135 | 0.13 |
| 3 | 0.675 | 0.717 | 0.639 | 0.857 | 0.738 | 0.664 | 0.622 | 0.407 | 0.593 | 0.501 | 0.642 | 0.452 |
| 4 | 0.767 | 0.942 | 1.111 | 0.819 | 0.784 | 0.808 | 0.636 | 0.702 | 0.802 | 0.567 | 0.63 | 0.665 |
| 5 | 0.791 | 0.957 | 0.943 | 0.705 | 0.415 | 0.641 | 0.615 | 0.609 | 0.835 | 0.609 | 0.592 | 0.488 |
| 6 | 0.868 | 1.075 | 0.644 | 0.557 | 0.509 | 0.687 | 0.429 | 0.476 | 0.601 | 0.588 | 0.487 | 0.712 |
| 7 | 0.368 | 1.056 | 0.87 | 0.551 | 0.604 | 0.693 | 0.47 | 0.466 | 0.577 | 0.559 | 0.723 | 0.94 |
| 8 | 0.786 | 1.05 | 0.933 | 0.829 | 0.723 | 0.795 | 0.649 | 0.588 | 0.767 | 0.638 | 0.715 | 0.716 |
| 9 | $\begin{aligned} & 0.786 \\ & \text { year } \end{aligned}$ | 1.05 | 0.933 | 0.829 | 0.723 | 0.795 | 0.649 | 0.588 | 0.767 | 0.638 | 0.715 | 0.716 |
| age | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 2 | 0.253 | 0.18 | 0.174 | 0.178 | 0.233 | 0.26 | 0.103 | 0.159 | 0.154 | 0.146 | 0.112 | 0.0907 |
| 3 | 0.693 | 0.578 | 0.597 | 0.408 | 0.393 | 0.636 | 0.332 | 0.343 | 0.34 | 0.304 | 0.397 | 0.3277 |
| 4 | 1.013 | 0.779 | 0.629 | 0.606 | 0.516 | 0.724 | 0.552 | 0.433 | 0.527 | 0.386 | 0.595 | 0.437 |
| 5 | 0.664 | 0.752 | 0.59 | 0.543 | 0.488 | 0.765 | 0.505 | 0.501 | 0.441 | 0.48 | 0.587 | 0.4642 |
| 6 | 0.887 | 0.9 | 0.502 | 0.679 | 0.407 | 0.755 | 0.756 | 0.686 | 0.45 | 0.361 | 0.439 | 0.5583 |
| 7 | 0.543 | 0.645 | 0.637 | 0.574 | 0.561 | 0.637 | 0.757 | 0.823 | 0.658 | 0.299 | 0.473 | 0.7306 |
| 8 | 0.853 | 0.808 | 0.682 | 0.616 | 0.531 | 0.795 | 0.623 | 0.606 | 0.533 | 0.409 | 0.555 | 0.5508 |
| 9 | $\begin{aligned} & 0.853 \\ & \text { year } \end{aligned}$ | 0.808 | 0.682 | 0.616 | 0.531 | 0.795 | 0.623 | 0.606 | 0.533 | 0.409 | 0.555 | 0.5508 |
| age | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |
| 2 | 0.121 | 0.0667 | 0.0378 | 0.0375 | 0.0341 | 0.022 | 0.024 | 0.0161 | 0.0111 | 0.0074 | 0.0052 |  |
| 3 | 0.265 | 0.2377 | 0.1935 | 0.1918 | 0.1744 | 0.112 | 0.123 | 0.0825 | 0.0568 | 0.0377 | 0.0265 |  |
| 4 | 0.393 | 0.3444 | 0.3886 | 0.3851 | 0.3502 | 0.226 | 0.246 | 0.1656 | 0.114 | 0.0758 | 0.0532 |  |
| 5 | 0.543 | 0.3392 | 0.3605 | 0.3573 | 0.3249 | 0.209 | 0.229 | 0.1536 | 0.1058 | 0.0703 | 0.0494 |  |
| 6 | 0.364 | 0.2856 | 0.4022 | 0.3986 | 0.3625 | 0.234 | 0.255 | 0.1714 | 0.118 | 0.0784 | 0.0551 |  |
| 7 | 0.427 | 0.4888 | 0.3405 | 0.3375 | 0.3069 | 0.198 | 0.216 | 0.1451 | 0.0999 | 0.0664 | 0.0466 |  |
| 8 | 0.436 | 0.375 | 0.3605 | 0.3573 | 0.3249 | 0.209 | 0.229 | 0.1536 | 0.1058 | 0.0703 | 0.0494 |  |
| 9 | 0.436 | 0.375 | 0.3605 | 0.3573 | 0.3249 | 0.209 | 0.229 | 0.1536 | 0.1058 | 0.0703 | 0.0494 |  |

Table 6.7.3.4. Irish Sea plaice: Update ICA stock summary.

| Year |  | Recruitment <br> Age 2 | TSB | SSB | Fbar <br> (Ages 3-6) | Landings | $\begin{aligned} & \text { Landings } \\ & \text { SOP } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | thousands | Tonnes | Tonnes | $f$ | Tonnes |  |
|  | 1964 | 21672 | 11755 | 8302 | 0.3031 | 2879 | 1 |
|  | 1965 | 29108 | 14182 | 9385 | 0.3644 | 3664 | 1 |
|  | 1966 | 15094 | 13805 | 9910 | 0.4248 | 4268 | 1 |
|  | 1967 | 13596 | 13277 | 10116 | 0.5067 | 5059 | 1 |
|  | 1968 | 10907 | 12109 | 9586 | 0.4797 | 4695 | 0.999 |
|  | 1969 | 12508 | 11275 | 9041 | 0.4605 | 4394 | 0.999 |
|  | 1970 | 18548 | 10928 | 8328 | 0.402 | 3583 | 1 |
|  | 1971 | 17283 | 10897 | 8153 | 0.6395 | 4232 | 1 |
|  | 1972 | 11825 | 11846 | 8977 | 0.6115 | 5119 | 1 |
|  | 1973 | 8746 | 9190 | 7027 | 0.7653 | 5060 | 1 |
|  | 1974 | 11710 | 7348 | 5437 | 0.7689 | 3715 | 1 |
|  | 1975 | 11514 | 7902 | 5680 | 0.7755 | 4063 | 0.999 |
|  | 1976 | 9664 | 5600 | 3909 | 0.9228 | 3473 | 0.999 |
|  | 1977 | 15014 | 4818 | 3003 | 0.8343 | 2904 | 0.989 |
|  | 1978 | 16168 | 5992 | 3541 | 0.7346 | 3231 | 1 |
|  | 1979 | 20089 | 7097 | 4154 | 0.6115 | 3428 | 1 |
|  | 1980 | 18194 | 7690 | 4625 | 0.6999 | 3903 | 1 |
|  | 1981 | 13554 | 8314 | 5417 | 0.5756 | 3906 | 0.999 |
|  | 1982 | 7389 | 7199 | 5154 | 0.5487 | 3237 | 0.999 |
|  | 1983 | 18737 | 6869 | 4543 | 0.7076 | 3639 | 0.996 |
|  | 1984 | 18777 | 8575 | 5468 | 0.5665 | 4241 | 1.006 |
|  | 1985 | 19800 | 10025 | 6374 | 0.5877 | 5075 | 1.002 |
|  | 1986 | 14385 | 10881 | 7161 | 0.5791 | 4806 | 1.011 |
|  | 1987 | 17600 | 10201 | 6860 | 0.8145 | 6220 | 1.001 |
|  | 1988 | 18827 | 13058 | 7350 | 0.7523 | 5005 | 0.997 |
|  | 1989 | 11496 | 11273 | 6709 | 0.5795 | 4372 | 0.996 |
|  | 1990 | 6605 | 8660 | 5760 | 0.5588 | 3275 | 0.998 |
|  | 1991 | 10234 | 7875 | 4806 | 0.4509 | 2554 | 0.984 |
|  | 1992 | 9041 | 7088 | 4697 | 0.72 | 3267 | 0.99 |
|  | 1993 | 9897 | 7040 | 4021 | 0.5362 | 1996 | 0.998 |
|  | 1994 | 8267 | 6217 | 3929 | 0.4907 | 2066 | 0.993 |
|  | 1995 | 7147 | 6039 | 3786 | 0.4395 | 1874 | 0.998 |
|  | 1996 | 6691 | 5809 | 4003 | 0.3826 | 1707 | 0.997 |
|  | 1997 | 8302 | 5466 | 3700 | 0.5047 | 1871 | 0.998 |
|  | 1998 | 8453 | 6156 | 3875 | 0.4468 | 1765 | 0.998 |
|  | 1999 | 7474 | 5924 | 3789 | 0.3913 | 1600 | 0.992 |
|  | 2000 | 7399 | 6119 | 3953 | 0.3017 | 1371 | 1 |
|  | 2001 | 9360 | 6817 | 4558 | 0.3362 | 1473 | 0.997 |
|  | 2002 | 7970 | 7405 | 4911 | 0.3332 | 1623 | 0.999 |
|  | 2003 | 9281 | 8471 | 5571 | 0.303 | 1559 | 1 |
|  | 2004 | 7777 | 7994 | 5422 | 0.1952 | 1143 | 0.997 |
|  | 2005 | 10195 | 8598 | 5903 | 0.2131 | 1281 | 1 |
|  | 2006 | 8835 | 9487 | 6592 | 0.1433 | 934 | 0.997 |
|  | 2007 | 8743 | 9490 | 6802 | 0.0986 | 805 | 0.998 |
|  | 2008 | 11666 | 12611 | 8821 | 0.0655 | 563 | 0.89 |
|  | 2009 | 3223 | 9647 | 7872 | 0.046 | 456 | 0.998 |

Table 6.7.4.1. Irish Sea plaice: input to short-term forecast for update run.
MFDP version 1a
Run: wgcse10
Time and date: 13:53 10/05/2010
Fbar age range: 3-6

| 2010 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | $N$ | $M$ | Mat | $P F$ | $P M$ | SWt | Sel | CWt |
| 2 | 8355 | 0.12 | 0.24 | 0 | 0 | 0.201 | 0.008 | 0.214 |
| 3 | 7077 | 0.12 | 0.57 | 0 | 0 | 0.234 | 0.040 | 0.245 |
| 4 | 8871 | 0.12 | 0.74 | 0 | 0 | 0.271 | 0.081 | 0.280 |
| 5 | 5508 | 0.12 | 0.93 | 0 | 0 | 0.311 | 0.075 | 0.319 |
| 6 | 4485 | 0.12 | 1 | 0 | 0 | 0.356 | 0.084 | 0.362 |
| 7 | 3959 | 0.12 | 1 | 0 | 0 | 0.404 | 0.071 | 0.408 |
| 8 | 2204 | 0.12 | 1 | 0 | 0 | 0.456 | 0.075 | 0.459 |
| 9 | 2710 | 0.12 | 1 | 0 | 0 | 0.557 | 0.075 | 0.557 |
|  |  |  |  |  |  |  |  |  |
| 2011 |  |  |  |  |  |  |  |  |
| Age | $N$ | $M$ | Mat | $P F$ | $P M$ | SWt | Sel | CWt |
| 2 | 8355 | 0.12 | 0.24 | 0 | 0 | 0.201 | 0.008 | 0.214 |
| 3 |  | 0.12 | 0.57 | 0 | 0 | 0.234 | 0.040 | 0.245 |
| 4 |  | 0.12 | 0.74 | 0 | 0 | 0.271 | 0.081 | 0.280 |
| 5 |  | 0.12 | 0.93 | 0 | 0 | 0.311 | 0.075 | 0.319 |
| 6 |  | 0.12 | 1 | 0 | 0 | 0.356 | 0.084 | 0.362 |
| 7 |  | 0.12 | 1 | 0 | 0 | 0.404 | 0.071 | 0.408 |
| 8 |  | 0.12 | 1 | 0 | 0 | 0.456 | 0.075 | 0.459 |
| 9 |  | 0.12 | 1 | 0 | 0 | 0.557 | 0.075 | 0.557 |


| 2012 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | $N$ | $M$ | Mat | $P F$ | $P M$ | SWt | Sel | CWt |
| 2 | 8355 | 0.12 | 0.24 | 0 | 0 | 0.201 | 0.008 | 0.214 |
| 3 |  | 0.12 | 0.57 | 0 | 0 | 0.234 | 0.040 | 0.245 |
| 4 |  | 0.12 | 0.74 | 0 | 0 | 0.271 | 0.081 | 0.280 |
| 5 | 0.12 | 0.93 | 0 | 0 | 0.311 | 0.075 | 0.319 |  |
| 6 | 0.12 | 1 | 0 | 0 | 0.356 | 0.084 | 0.362 |  |
| 7 |  | 0.12 | 1 | 0 | 0 | 0.404 | 0.071 | 0.408 |
| 8 | 0.12 | 1 | 0 | 0 | 0.456 | 0.075 | 0.459 |  |
| 9 | 0.12 | 1 | 0 | 0 | 0.557 | 0.075 | 0.557 |  |

Input units are thousands and kg - output in tonnes

Table 6.7.4.2. Irish Sea plaice: Single option prediction detailed forecast for update run.

MFDP version 1a
Run: wgcse10
Time and date: $13: 53$ 10/05/2010
Fbar age range: 3-6

| Year: | 2010 | F multiplier: 1 |  | Fbar: | 0.0701 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 | 0.0079 | 62 | 13 | 8355 | 1679 | 2005 | 403 | 2005 | 403 |
| 3 | 0.0403 | 264 | 65 | 7077 | 1658 | 4034 | 945 | 4034 | 945 |
| 4 | 0.081 | 651 | 182 | 8871 | 2404 | 6565 | 1779 | 6565 | 1779 |
| 5 | 0.0751 | 376 | 120 | 5508 | 1713 | 5122 | 1593 | 5122 | 1593 |
| 6 | 0.0838 | 340 | 123 | 4485 | 1595 | 4485 | 1595 | 4485 | 1595 |
| 7 | 0.071 | 256 | 104 | 3959 | 1598 | 3959 | 1598 | 3959 | 1598 |
| 8 | 0.0751 | 150 | 69 | 2204 | 1005 | 2204 | 1005 | 2204 | 1005 |
| 9 | 0.0751 | 185 | 103 | 2710 | 1509 | 2710 | 1509 | 2710 | 1509 |
| Total |  | 2284 | 779 | 43169 | 13162 | 31084 | 10427 | 31084 | 10427 |
| Year: | 2011 | F multiplier: 1 |  | Fbar: | 0.0701 |  |  |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 | 0.0079 | 62 | 13 | 8355 | 1679 | 2005 | 403 | 2005 | 403 |
| 3 | 0.0403 | 274 | 67 | 7352 | 1723 | 4191 | 982 | 4191 | 982 |
| 4 | 0.081 | 442 | 124 | 6028 | 1634 | 4461 | 1209 | 4461 | 1209 |
| 5 | 0.0751 | 495 | 158 | 7256 | 2257 | 6748 | 2099 | 6748 | 2099 |
| 6 | 0.0838 | 344 | 124 | 4532 | 1612 | 4532 | 1612 | 4532 | 1612 |
| 7 | 0.071 | 236 | 96 | 3658 | 1477 | 3658 | 1477 | 3658 | 1477 |
| 8 | 0.0751 | 223 | 103 | 3271 | 1491 | 3271 | 1491 | 3271 | 1491 |
| 9 | 0.0751 | 276 | 154 | 4043 | 2251 | 4043 | 2251 | 4043 | 2251 |
| Total |  | 2353 | 839 | 44494 | 14123 | 32908 | 11523 | 32908 | 11523 |


| Year: | 2012 | F multiplier: 1 | Fbar: | 0.0701 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 | 0.0079 | 62 | 13 | 8355 | 1679 | 2005 | 403 | 2005 | 403 |
| 3 | 0.0403 | 274 | 67 | 7352 | 1723 | 4191 | 982 | 4191 | 982 |
| 4 | 0.081 | 459 | 129 | 6263 | 1697 | 4635 | 1256 | 4635 | 1256 |
| 5 | 0.0751 | 337 | 107 | 4931 | 1533 | 4586 | 1426 | 4586 | 1426 |
| 6 | 0.0838 | 453 | 164 | 5969 | 2123 | 5969 | 2123 | 5969 | 2123 |
| 7 | 0.071 | 239 | 98 | 3696 | 1492 | 3696 | 1492 | 3696 | 1492 |
| 8 | 0.0751 | 206 | 95 | 3022 | 1378 | 3022 | 1378 | 3022 | 1378 |
| 9 | 0.0751 | 411 | 229 | 6017 | 3349 | 6017 | 3349 | 6017 | 3349 |
| Total |  | 2440 | 901 | 45605 | 14975 | 34120 | 12410 | 34120 | 12410 |

Input units are thousands and kg - output in tonnes

Table 6.7.4.3. Irish Sea plaice: Prediction with management options for update run.
MFDP version 1a
Run: wgcse10
IRISH SEA PLAICE, 2010 WG, Forecast Inputs
Time and date: 13:53 10/05/2010
Fbar age range: 3-6

| 2010 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 13162 | 10427 | 1.0000 | 0.0701 | 779 |  |  |
| 2011 |  |  |  |  | 2012 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 14123 | 11523 | 0.0 | 0.0000 | 0 | 15847 | 13248 |
|  | 11523 | 0.1 | 0.0070 | 87 | 15757 | 13162 |
|  | 11523 | 0.2 | 0.0140 | 173 | 15668 | 13076 |
| . | 11523 | 0.3 | 0.0210 | 258 | 15579 | 12990 |
|  | 11523 | 0.4 | 0.0280 | 343 | 15491 | 12905 |
| . | 11523 | 0.5 | 0.0350 | 427 | 15403 | 12821 |
|  | 11523 | 0.6 | 0.0420 | 510 | 15317 | 12738 |
|  | 11523 | 0.7 | 0.0491 | 593 | 15230 | 12655 |
|  | 11523 | 0.8 | 0.0561 | 676 | 15145 | 12573 |
|  | 11523 | 0.9 | 0.0631 | 758 | 15060 | 12491 |
| . | 11523 | 1.0 | 0.0701 | 839 | 14975 | 12410 |
| - | 11523 | 1.1 | 0.0771 | 919 | 14892 | 12329 |
| . | 11523 | 1.2 | 0.0841 | 1000 | 14809 | 12249 |
|  | 11523 | 1.3 | 0.0911 | 1079 | 14726 | 12170 |
|  | 11523 | 1.4 | 0.0981 | 1158 | 14644 | 12091 |
| . | 11523 | 1.5 | 0.1051 | 1236 | 14563 | 12013 |
|  | 11523 | 1.6 | 0.1121 | 1314 | 14482 | 11936 |
| . | 11523 | 1.7 | 0.1191 | 1391 | 14402 | 11859 |
|  | 11523 | 1.8 | 0.1261 | 1468 | 14322 | 11782 |
| . | 11523 | 1.9 | 0.1331 | 1544 | 14243 | 11706 |
| . | 11523 | 2.0 | 0.1401 | 1620 | 14165 | 11631 |
| - | 11523 | 6.4 | 0.4485 | 4469 | 11221 | 8811 |
| - | 11523 | 6.42 | 0.4500 | 4480 | 11210 | 8800 |
| . | 11523 | 6.5 | 0.4555 | 4524 | 11164 | 8756 |

Input units are thousands and kg - output in tonnes
 weight) of these year classes.

| Year-class |  |  | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock No. (thousands) |  |  | 8743 | 11666 | 8355 | 8355 | 8355 |
| of Source 2 year-olds |  |  |  |  |  |  |  |
|  |  |  | ICA | ICA | GM90-07 | GM90-07 | GM90-07 |
| Status Quo F: |  |  |  |  |  |  |  |
| \% in | 2010 | landings | 16.5 | 28.5 | 11.6 | 2.7 | - |
| \% in | 2011 |  | 14.6 | 21.0 | 18.8 | 11.6 | 2.6 |
| \% in | 2010 | SSB | 15.3 | 17.1 | 9.1 | 3.9 | - |
| \% in | 2011 | SSB | 14.0 | 18.2 | 10.5 | 8.5 | 3.5 |
| \% in | 2012 | SSB | 12.0 | 17.1 | 11.5 | 10.1 | 7.9 |

GM : geometric mean recruitment
Plaice in Vlla - Final run. : Year-class \% contribution to


Table 6.7.4.5. Irish Sea plaice: Final run - Yield-per-recruit table under current selection pattern.
MFYPR version 2a
Run: wgcse2010
Time and date: 14:13 10/05/2010

| Yield per results <br> FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 8.8433 | 3.6278 | 7.4486 | 3.3150 | 7.4486 |  |
| 0.1000 | 0.0070 | 0.0509 | 0.0225 | 8.4196 | 3.4021 | 7.0268 | 3.0898 | 7.0268 |  |
| 0.2000 | 0.0140 | 0.0962 | 0.0419 | 8.0431 | 3.2024 | 6.6522 | 2.8907 | 6.6522 | 2.0898 |
| 0.3000 | 0.0210 | 0.1366 | 0.0589 | 7.7063 | 3.0246 | 6.3173 | 2.7135 | 6.3173 | 2.7907 |
| 0.4000 | 0.0280 | 0.1731 | 0.0738 | 7.4033 | 2.8655 | 6.0162 | 2.5548 | 6.0162 | 2.5548 |
| 0.5000 | 0.0350 | 0.2060 | 0.0870 | 7.1292 | 2.7222 | 5.7439 | 2.4121 | 5.7439 | 2.4121 |
| 0.6000 | 0.0420 | 0.2360 | 0.0986 | 6.8801 | 2.5927 | 5.4967 | 2.2831 | 5.4967 | 2.2831 |
| 0.7000 | 0.0491 | 0.2633 | 0.1089 | 6.6527 | 2.4750 | 5.2711 | 2.1659 | 5.2711 | 2.1659 |
| 0.8000 | 0.0561 | 0.2884 | 0.1181 | 6.4443 | 2.3677 | 5.0645 | 2.0592 | 5.0645 | 2.0592 |
| 0.9000 | 0.0631 | 0.3115 | 0.1263 | 6.2526 | 2.2696 | 4.8746 | 1.9615 | 4.8746 | 1.9615 |
| 1.0000 | 0.0701 | 0.3328 | 0.1336 | 6.0757 | 2.1795 | 4.6995 | 1.8719 | 4.6995 | 1.8719 |
| 1.1000 | 0.0771 | 0.3525 | 0.1402 | 5.9119 | 2.0966 | 4.5375 | 1.7895 | 4.5375 | 1.7895 |
| 1.2000 | 0.0841 | 0.3708 | 0.1462 | 5.7598 | 2.0200 | 4.3872 | 1.7134 | 4.3872 | 1.7134 |
| 1.3000 | 0.0911 | 0.3878 | 0.1516 | 5.6183 | 1.9491 | 4.2474 | 1.6430 | 4.2474 | 1.6430 |
| 1.4000 | 0.0981 | 0.4037 | 0.1564 | 5.4861 | 1.8833 | 4.1170 | 1.5777 | 4.1170 | 1.5777 |
| 1.5000 | 0.1051 | 0.4186 | 0.1608 | 5.3626 | 1.8221 | 3.9952 | 1.5170 | 3.9952 | 1.5170 |
| 1.6000 | 0.1121 | 0.4326 | 0.1648 | 5.2467 | 1.7651 | 3.8811 | 1.4604 | 3.8811 | 1.4604 |
| 1.7000 | 0.1191 | 0.4457 | 0.1684 | 5.1379 | 1.7118 | 3.7740 | 1.4076 | 3.7740 | 1.4076 |
| 1.8000 | 0.1261 | 0.4581 | 0.1717 | 5.0355 | 1.6620 | 3.6733 | 1.3583 | 3.6733 | 1.3583 |
| 1.9000 | 0.1331 | 0.4697 | 0.1747 | 4.9390 | 1.6153 | 3.5785 | 1.3120 | 3.5785 | 1.3120 |
| 2.0000 | 0.1401 | 0.4807 | 0.1774 | 4.8478 | 1.5714 | 3.4890 | 1.2686 | 3.4890 | 1.2686 |


| Reference point | F multiplier Absolute F |  |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.0701 |
| FMax | 8.9160 | 0.6248 |
| F0.1 | 2.0861 | 0.1462 |
| F35\%SPR | 2.2801 | 0.1598 |

Weights in kilograms

Table 6.7.4.6. Irish Sea plaice: Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships (Ricker, Beverton-Holt, smooth hockeystick) and the yield-per-recruit $F_{\text {msy }}$ proxies.

| Ricker |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 525/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AlCe |
| Deterministic | 1.30379 | 0.360702 | 7361.88 | 2462.83 | 0.701422 | 1.06482 | 4.17742 | 0.000128261 | 48.1924 |
| Mean | 1.446129017 | 0.358935785 | 8385.039619 | 2693.954324 | 0.713836411 | 1.034744785 | 4.178940076 | 0.000124638 |  |
| 5\%ile | 0.7114112 | 0.243361 | 5529.66 | 1986.168 | 0.6020188 | 0.6247296 | 2.97412 | $7.53 \mathrm{E}-05$ |  |
| 25\%ile | 0.938627 | 0.2983 | 6690.01 | 2343.96 | 0.66573 | 0.866909 | 3.60577 | 0.000104422 |  |
| 50\%ile | 1.2196 | 0.351278 | 7832.42 | 2640.56 | 0.711289 | 1.0285 | 4.13509 | 0.000123886 |  |
| 75\%ile | 1.76443 | 0.412307 | 9212.08 | 2910.28 | 0.757621 | 1.20384 | 4.66155 | 0.000145007 |  |
| 95\%ile | 2.891902 | 0.5046566 | 11850.72 | 3428.244 | 0.822159 | 1.432274 | 5.50803 | 0.000172521 |  |
| CV | 0.490664218 | 0.229907819 | 0.418691736 | 0.253501936 | 0.097567486 | 0.245231795 | 0.188086177 | 0.245231836 |  |
| Beverton-Holt |  |  |  |  |  |  |  |  |  |
| 225/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AlCe |
| Deterministic | 5 | 0.283499 | 8765.86 | 2367.74 | 1.30645 | 1.46605 | 13048.8 | 1014.13 | 46.3924 |
| Mean | 2.339334169 | 0.181042194 | 33704.97138 | 3121.248978 | 0.995923169 | 1.369328844 | 18178.34533 | 3613.132253 |  |
| 5\%ile | 0.8383842 | 0.0242317 | 7800.28 | 2030.506 | 0.6707234 | 1.216166 | 13191.84 | 1259.238 |  |
| 25\%ile | 1.3744 | 0.130761 | 11131.7 | 2448.24 | 0.863055 | 1.30655 | 14878.9 | 1978.53 |  |
| 50\%ile | 2.19428 | 0.195183 | 14891.6 | 2860.55 | 1.00291 | 1.36861 | 16998.3 | 2982.8 |  |
| 75\%ile | 3.06495 | 0.237183 | 24274.9 | 3432.61 | 1.14576 | 1.43992 | 19752.7 | 4316.67 |  |
| 95\%ile | 4.278444 | 0.290122 | 171687.4 | 4907.272 | 1.292314 | 1.532432 | 25416.86 | 7636.952 |  |
| CV | 0.474484283 | 0.43278769 | 1.584233421 | 0.37087491 | 0.200848915 | 0.074055602 | 0.361308292 | 0.864867062 |  |
| Smooth hockeystick |  |  |  |  |  |  |  |  |  |
| 542/1000 Iterations resulted in feasible parameter estimates |  |  |  |  |  |  |  |  |  |
|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AlCC |
| Deterministic | 0.668061 | 0.62446 | 4486.07 | 2357.1 | 0.647442 | 0.707805 | 1.32949 | 4261.92 | 49.6155 |
| Mean | 0.850218749 | 0.456712244 | 16039.79068 | 2523.126328 | 0.739977157 | 0.638760079 | 1.519501764 | 3846.182638 |  |
| 5\%ile | 0.50842485 | 0.028699305 | 3278.253 | 1965.195 | 0.55461095 | 0.50552535 | 1.13886 | 3043.9265 |  |
| 25\%ile | 0.656644 | 0.291264 | 4490.9925 | 2246.47 | 0.64905875 | 0.5386975 | 1.3328025 | 3243.67 |  |
| 50\%ile | 0.7842695 | 0.44772 | 6145.725 | 2437.365 | 0.7487845 | 0.5930465 | 1.537585 | 3570.925 |  |
| 75\%ile | 0.958064 | 0.596283 | 9407.2525 | 2729.1625 | 0.83479925 | 0.7153345 | 1.714215 | 4307.265 |  |
| 95\%ile | 1.4041595 | 0.9454021 | 87229.955 | 3367.233 | 0.91568425 | 0.87458945 | 1.880306 | 5266.184 |  |
| CV | 0.363134893 | 0.588311531 | 1.939550863 | 0.171933832 | 0.154600765 | 0.216403775 | 0.154600775 | 0.216403777 |  |
| Per recruit |  |  |  |  |  |  |  |  |  |
|  | F35 | F40 | F01 | Fmax | Bmsypr | MSYpr | Fpa | Flim |  |
| Deterministic | 0.159733 | 0.130935 | 0.146148 | 0.625148 | 0.395865 | 0.207998 |  | 0.45 |  |
| Mean | 0.144440731 | 0.118549045 | 0.129051199 | 0.670015642 | 1.408993089 | 0.222273389 | 49.95124926 |  |  |
| 5\%ile | 0.001728576 | 0.001396494 | 0.002001092 | 0.03256357 | 0.2954358 | 0.17576805 | 47.430855 |  |  |
| 25\%ile | 0.07597555 | 0.06195255 | 0.0710139 | 0.30744425 | 0.39264 | 0.20065825 | 48.41685 |  |  |
| 50\%ile | 0.1528285 | 0.12593 | 0.1403895 | 0.467667 | 0.542314 | 0.217633 | 49.73625 |  |  |
| 75\%ile | 0.214397 | 0.176807 | 0.18625625 | 0.77163125 | 0.8126365 | 0.2371355 | 50.849725 |  |  |
| 95\%ile | 0.2841279 | 0.235097 | 0.23026565 | 2.120171 | 7.785563 | 0.2836772 | 53.761225 |  |  |
| CV | 0.628068054 | 0.627919852 | 0.581019904 | 0.924230211 | 1.92542654 | 0.152337271 | 0.044153159 |  |  |

Effort for beam trawls (000 hours)


Figure 6.7.2.1. Irish Sea plaice: Effort and lpue for commercial fleets.


Figure 6.7.2.2. Landings and survey data. Raw landings-at-age data (top left), mean standardised proportion-at-age (topright, grey bubbles are positive values and white bubbles are negative), UK (E\&W) beam trawl survey cpue for ages 1 to 5 , and mean standardised indices of spawning biomass (bottom right) derived from UK (NI) groundfish surveys (UK(NI)GFS) in March and in October and, for comparison, the biomass index calculated from UK (E\&W) beam trawl survey in September.


Figure 6.7.2.3. Length distributions of discarded and retained catches from UK (E\&W).







Figure 6.7.2.4. Length distributions of discarded and retained catches from Ireland.


Figure 6.7.2.5. Length distributions of discarded and retained catches from Belgium.

UK BT SURVEY (Sept-Trad) - Prime stations only


Figure 6.7.2.6. Log cpue plot of UK BT survey by year.

UK BT SURVEY (Sept-Trad) - Prime stations only


Figure 6.7.2.7. Log cpue plot of UK BT survey by year-class.


Figure 6.7.2.8. Northern Irish groundfish survey SSB indices split into spring (left hand panels) and autumn (right hand panels) sampling by western strata (1-3), eastern strata (4-7) and total survey area (strata 1-7) with confidence intervals ( $\pm 1$ standard error, vertical lines) and mean biomass ( $\mathrm{kg} / 3 \mathrm{miles}$, dashed horizontal lines) for periods identified by statistical breakpoint analysis.

Note the different scale on the $y$-axis in the top-left panel.


Figure 6.7.3.1. ICA residuals for UK (E\&W) beam trawl survey (BTS) at age (top two rows), UK (NI)GFS SSB indices (middle 2 panes) and separable model residuals (below).


Figure 6.7.3.2 Retrospective pattern for update ICA.


Figure 6.7.3.3. Irish Sea plaice: Summary plot for update ICA assessment.


Figure 6.7.3.4. Comparison of recruitment (age 2), SSB and Fbar(ages 3-6) between 2009 and 2010 ICA assessments.


| MFYPR version 2a |  |  |
| :--- | :--- | :--- |
| Run: wgcse2010 |  |  |
| Time and date: 14:13 | 10/05/2010 |  |
| Reference point | F multiplier | Absolute F |
| Fbar(3-6) | 1.0000 | 0.0701 |
| FMax | 8.9160 | 0.6248 |
| F0.1 | 2.0861 | 0.1462 |
| F35\%SPR | 2.2801 | 0.1598 |
|  |  |  |
| Weights in kilograms |  |  |

MFDP version 1
Run: wgcse10
IRISH SEA PLAICE, 2010 WG, Forecast Inputs
Time and date: 13:53 10/05/2010
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Figure 6.7.3.5. VIIa plaice, yield-per-recruit and short-term forecast from final ICA


Figure 6.7.3.6. Plaice in Division VIIa : MSY fitted stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $\mathbf{F}_{\text {msy. }}$. Right hand panels: curves plotted from the first 100 MCMC resamples with converged $\mathrm{F}_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of Fmsy from the $\mathbf{1 0 0 0}$ re-samples.
wgcse_ple-iris Ricker


Figure 6.7.3.7a. Plaice in Division VIIa: Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Ricker stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $F_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $F_{\text {msy }}$ estimates. Circles in left hand panels show assessment estimates with the most recent year labelled.


Figure 6.7.3.7b. Plaice in Division VIIa: Estimates of F reference points and equilibrium yield and SSB against fishing mortality using a Smooth hockeystick model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathbf{F}_{\text {msy. }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles in left hand panels show assessment estimates with the most recent year labelled.


Figure 6.7.3.8. Plaice in Division VIIa: Fitted yield-per-recruit F reference points, yield-per-recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first 100 re-samples.

### 6.8 Sole in Division VIIa (Irish Sea)

Type of assessment in 2010
This assessment is an Update Assessment, and consequently no changes have been made to assessment or forecast procedures compared to last year.

## ICES advice applicable to 2009

"Given the low SSB and low recruitment since 2000, it is not possible to identify any nonzero catch which would be compatible with the precautionary approach. ICES recommends a closure of the fishery in 2009 and a recovery plan should be developed and implemented as a prerequisite to reopening the fishery"

## ICES advice applicable to 2010

"Considering the options below, ICES advises on the basis of exploitation boundaries in relation to precautionary limits that no fishing of sole should take place in the Irish Sea in 2010."

### 6.8.1 General

## Stock description and management units

The sole fisheries in the Irish Sea are managed by TAC (see text tables below) and technical measures, with the assessment area corresponding to the stock area. Technical measures in force are minimum mesh sizes and minimum landing size $(24 \mathrm{~cm})$. In addition beam trawlers, fishing with mesh sizes equal to or greater than 80 mm , are obliged to have 180 mm mesh sizes in the entire upper half of the anterior part of their net. More details can be found in Reg 254/2002 and the Stock Annex.

Since 2000, a spawning closure for cod has been in force. The first year of the regulation the closure covered the western and eastern Irish Sea. Since then, closure has been mainly in the western part whereas the sole fishery takes place mainly in the eastern part of the Irish Sea and no direct impact on the sole stock is expected from this closure.

Other regulations applicable to area VIIa are summarized in Section 6.1.

## Management applicable to 2009 and 2010

TAC 2009

| Species: Common sole <br> Solea solea |  | Zone:VIIa <br> (SOL/07A.) |  |
| :--- | ---: | ---: | :--- |
| Belgium | 237 |  |  |
| France | 3 |  |  |
| Ireland | 80 |  |  |
| The Netherlands | 75 | Analytical TAC <br> United Kingdom <br> EC | Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |
| TAC | 502 |  |  |

TAC 2010

| Species:Common sole <br> Solea solea | Zone:VIla <br> (SOL/07A.) |  |
| :--- | :--- | :--- | :--- |
| Belgium | 186 |  |
| France | 2 |  |
| Ireland | 73 |  |
| The Netherlands | 58 |  |
| United Kingdom | 83 |  |
| EU | 402 | Analytical TAC |
| TAC | 402 |  |


| Year | Single stock exploitation bounderies | Basis | TAC | \% change in <br> $F$ associated with TAC * | WG landings |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | $<1000$ t | Keep F below $\mathrm{F}_{\mathrm{pa}}$ | 960 t | +3 | 855 t |
| 2006 | $<930$ t | Recent catch levels (20022004) | 960 t | - | 570 t |
| 2007 | 0 t | Zero catch | 816 t | +1 | 492 t |
| 2008 | 0 t | Zero catch | 669 t | 0 | 333 t |
| 2009 | 0 t | Zero catch | 502 t | -3 |  |
| 2010 | 0 t | Zero catch | 402 t |  |  |

F calculated, based on a Status quo forecast.

## Fishery in 2009

The main countries fishing for Irish Sea sole are Belgium, Ireland and UK.
Effort declined in all the main fleets in 2009 with the exception of Belgium which increased.

### 6.8.2 Data

## Landings

An overview of the landings data provided and used by the WG is provided in Table 6.8.1. The WG estimated the total international landings at 324 t in 2009, of which $79 \%$ $(257 \mathrm{t})$ was landed by Belgium, $15 \%(47 \mathrm{t})$ by Ireland, $3 \%(10 \mathrm{t})$ by the UK (England and Wales) and the remainder by Northern Ireland and France (Table 6.8.1). These landing-figures correspond to an international uptake of only $65 \%$ of the agreed TAC in 2009 (502 t), and the lowest value in the time-series.

No revisions were made to the historical data.
Quarterly age compositions for 2009 were available from Belgium, UK (E\&W) (except Q1) and Ireland (except Q1 and Q2) as well as quarterly landings from Northern Ireland. The sampled fleets are those taking the major part of the international landings. Annual length distributions are given in Table 6.8.2.

Catch numbers-at-age data are given in Table 6.8.3.
Catch weights-at-age for 2009 were calculated from Belgian, UK and Irish data. Stock weights-at-age were derived from the Q1 catch weights (Table 6.8.5).

Further details on raising methods are given in the Stock Annex.

## Discards

Discard rates of sole are low in Irish Sea fisheries based on historical observation. No discard tables and figures are presented in this report.

There is no accurate information on the level of misreporting, but given the partial uptake of the agreed TAC in recent years, it is not considered a problem for this stock.

## Biological

Natural mortality, maturity and proportions of natural mortality and fishing mortality before spawning were set as in previous years.

Natural mortality was set at $0.1 \mathrm{yr}^{-1}$ (all ages and all years).
The maturity ogive used is as previously:

| AGE | $\mathbf{1}$ |  | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ AND OLDER |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.00 |  | 0.38 |  | 0.71 |  | 0.97 |  |

The proportions of natural mortality and fishing mortality before spawning were both set to 0 to reflect the SSB calculation date of 1 January.

## Surveys

Cpue and effort-series were available from a UK (E\&W) September beam trawl survey (1988-2009) and a UK March beam trawl survey (1993-1999) (Tables 6.8.6-7 and Figure 6.8.1). From 2006 onwards, only the two UK beam trawl surveys have been used as tuning indices in the Irish Sea sole assessments.

## Commercial cpue

Commercial tuning data were available for Belgian beam trawlers, UK (E\&W) beam and otter trawlers and Irish otter trawlers (Tables 6.8.6-7 and Figure 6.8.1).

Effort from both Belgian and UK commercial beam trawl fleets increased from the early seventies until the late eighties. Since then UK beam trawl effort has declined. The Belgian beam trawl effort declined in the early nineties, increased the early 2000s and decreased again recently. Effort of the Irish beam trawl fleet has been declining since 2000. In 2008 and 2009 effort declined substantially for most fleets (Tables 6.8.67 and Figure 6.8.1).

Lpue for both UK and Belgian beam trawlers was at a higher level in the late seventies and early eighties. More recently lpue for these beam trawlers is fluctuating at a lower level. Irish beam trawl lpue also shows a declining trendover the time-series.

### 6.8.3 Historical stock development

The method used to assess Irish Sea sole is XSA, using two survey tuning-series (Tables 6.8.6-7). It should be noted that the year range of the UK March beam trawl survey only covers 1993 up to 1999.

## Data screening

No exploratory runs using different settings than last year's assessment, or taking revisions from earlier data into account, are presented in this report for Irish Sea sole.

## Final update assessment

The model settings for the final assessment are summarized below. Since this is an update assessment, settings were kept the same as last year. Log catchability residuals for the final run are given in Figure 6.8.2. There are no apparent trends. The XSA diagnostics and the estimates of fishing mortality and the population numbers are given in Tables 8.6.8-10. The summary is given in Table 6.8.11 and Figure 6.8.3.

| Assmnt Year | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Assmnt <br> Model | : XSA | : XSA | : XSA | :XSA | :XSA |
| Fleets | : | : | : | : | : |
| Bel Beam Trwl | : omitted | : omitted | : omitted | :omitted | :omitted |
| UK Trawl | : omitted | : omitted | : omitted | :omitted | :omitted |
| UK Sept BTS | :1988-2005 | : 1988-2006 | : 1988-2007 | :1988-2008 | :1988-2008 |
|  | 2-7 | 2-7 | 2-7 | 2-7 | 2-7 |
| UK Mar BTS | 1993-1999 | : 1993-1999 | : 1993-1999 | :1993-1999 | :1993-1999 |
|  | 2-7 | 2-7 | 2-7 | 2-7 | 2-7 |
| Time-Ser. Wts | :linear 20 yrs | : linear 20 yrs | : linear 20 yrs | : linear 20 yrs | : linear 20 yrs |
| Power Model | :: none | : none | : none | : none | : none |
| Q plateau | :: 5 | : 7 | : 7 | : 7 | : 7 |
| Shk se | : 1.5 | : 1.5 | : 1.5 | :1.5 | :1.5 |
| Shk age-yr | :5 yrs 3 ages | : 5 yrs 3 ages | : 5 yrs 3 ages | : 5 yrs 3 ages | : 5 yrs 3 ages |
| Pop Shk se | : 0.3 | : 0.3 | : 0.3 | : 0.3 | : 0.3 |
| Prior Wting | : none | : none | : none | : none | : none |
| Plusgroup | : 10 | : 8 | : 8 | : 8 | : 8 |
| Fbar | : 4-7 | : 4-7 | : 4-7 | : 4-7 | : 4-7 |

Survivor and F estimates coming from the UK (E+W) September beam trawl survey and from F shrinkage are not always in line with each other. But given that the survey gets high weights ( $>96 \%$ ) throughout, the survey has a bigger influence on the final estimates. The March survey was discontinued after 1999, and therefore does not contribute to the estimates in the final year.

The retrospective analysis is presented in Figure 6.8.4. A retrospective pattern is apparent in SSB, although for most recent estimates the yearly revisions are minor. Recruitment levels appear to be consistently estimated throughout the retrospective period.

## Comparison with previous assessments

A comparison of the estimates of this year's assessment with last year's is given in Figure 6.8.5. Recruitment trends, historical SSB and fishing mortality estimates are very similar.

## State of the stock

Estimated trends of Irish Sea sole landings, SSB, fishing mortality and recruitment are presented in Table 6.8.11 and Figure 6.8.3. Landings of Irish Sea sole have been declining since the late eighties and reached a record low of 324 t in 2009. SSB has been at a lower level since the early nineties compared to the period before. Since 2001 SSB has been decreasing dramatically and reached the lowest observed estimate in 2008 (for the first time only reaching values below 2000 t since 2006). High fishing mortalities were observed during the late eighties until the mid-nineties. Thereafter fishing mortality has declined somewhat, but remained fluctuating around Flim. Since 2001 recruitment has been well below average.

### 6.8.4 Short-term projections

## Estimating year-class abundance

The estimates up to the 2006 year class were taken from XSA.
The 2007 year class (age 2 in 2009) was estimated using RCT3 (input in Table 6.8.12, output in Table 6.8.13). Both RCT3 and XSA estimate a weak 2005 year class (higher for RCT3), but the RCT3 estimate was taken over the XSA estimate since it uses more recent survey data and is in line with last year's procedure.

The different estimates at age 2 are summarized below. The values in bold were selected for further predictions.

| YEAR CLASS | $: 2007$ | $: 2008$ | $: 2009$ AND OLDER |
| :--- | :---: | :---: | :---: | :---: |
| XSA | $: 2397$ | $:-$ | $:-$ |
| RCT3 | $: 2877$ | $: 2489$ | $:-$ |
| GM | $: 5326$ | $: 5326$ | $: 5326$ |

The input to the short-term catch predictions is given in Table 6.8.14. Weights-at-age averaged over the last three years were used as input for the predictions. As for last year, fishing mortality-at-age was averaged over the last three years, not rescaled. XSA estimates up to year class 2007 were used for the starting population. For the year class 2008 the RCT3 estimate was used. GM over the full period was assumed for the recruiting ages from 2010 onwards.

The short-term catch option table is given in Table 6.8.15, a detailed management option table is presented in Table 6.8.16. A short-term forecast plot is shown in Figure 6.8.6. Assuming $\mathrm{F}_{\mathrm{sq}}$, landings in 2010 are estimated to be around 439 t , compared to a TAC of 402 t .

The relative contributions of the different year classes to the landings and SSB are presented in Table 6.8.17 and Figure 6.8.8. Given the low stock size, predictions become more dependent on the assumed incoming recruitment. $55 \%$ of the predicted landings in 2011 and $49 \%$ of the predicted SSB in 2012 are based on the assumed the RCT3 and GM recruitments.

### 6.8.5 MSY explorations

The VIIa sole times-series of assessment stock and recruit estimates, fishing mortality-at-age (average of the most recent three year), catch and stock weights (10 year averages), maturity and natural mortality-at-age were used to estimate proxies for the fishing mortality biomass and landings at maximum sustainable yield ( $\mathrm{F}_{\mathrm{msy}}, \mathrm{B}_{\mathrm{msy}}$ and MSY) within the srmsymc program. The sen and sum input data files are presented in Tables 6.8.19 and 6.8.20.

Three stock and recruit models are fitted by the program, Ricker, Beverton and Holt and the smooth hockey stock Figures 6.8.7-6.8.10. Based on the A.I.C. and number of successful fits all models have an equal fit to the available data. The estimates of $\mathrm{F}_{\text {crash, }}$, $\mathrm{F}_{\mathrm{msy}}, \mathrm{B}_{\mathrm{msy}}$ and MSY are presented with their percentiles and coefficients of variation in Table 6.8.21. Figure 6.8.11 illustrates the uncertainty in yield-per-recruit curve, with estimates also presented in Table 6.8.21.

Each model assumes that there is a relationship between increasing levels of recruitment and increasing spawning-stock. In the past recruitment seems to have show some cyclical pattern at stock sizes $>3000 \mathrm{t}$ suggesting the dominant determinant of recruitment at higher stock sixes is likely to be environmental conditions rather than the level of SSB.

The link to environmental control of recruitment and independence from SSB would suggest the use of yield-per-recruit fishing mortality reference levels as appropriate. However, as shown by Figure 6.8.11 the form of the YPR curve is poorly determined and the estimates of $\mathrm{F}_{0.1}, \mathrm{~F}_{35 / 40 \%} \mathrm{SPR}$ have high cv, with $\mathrm{F}_{\max }$ very poorly determined (Table 6.8.21).

For the Beverton and Holt model and the Ricker model $\mathrm{F}_{\text {msy }}$ overlaps with $\mathrm{F}_{0.1}$ and $\mathrm{F}_{35 \%}$ and has a high probability of avoiding $\mathrm{F}_{\text {crash. }}$. For the smooth hockeystick the breakpoint is poorly determined and far to the right resulting in an $\mathrm{F}_{\text {crash }}$ estimate that is close to $\mathrm{F}_{\text {msy, }} \mathrm{F}_{0.1}$ and $\mathrm{F}_{35 \%}$. Given that the stock has been fished above $\mathrm{F}_{\text {crash }}$ for the time-series these resulting F reference estimates from the smooth hockey stick were considered unlikely.

Estimates of Fmsy differ assumed B-H and Ricker stock and recruitment relationships are fitted to the S-R pairs for sole in VIIa, both models are equally plausible and there is no way of distinguishing between them. Consequently WGCSE consider that fishing mortalities in the range $0.07-0.16$ are consistent with maximising long-term yield for sole in VIIa.

### 6.8.6 Biological reference points

## Precautionary approach reference points

Biological reference points are:

| $\begin{aligned} & \mathrm{B} \lim =2200 \\ & \mathrm{t} \end{aligned}$ | Basis: $\mathrm{Blim}_{\text {lim }}=\mathrm{Bloss}^{\text {l }}$ | Changed in ACFM 2007 (from 2800 to 2200 t ). The lowest observed spawning stock, followed by an increase in SSB. |
| :---: | :---: | :---: |
| $\mathrm{B}_{\mathrm{pa}}=3100 \mathrm{t}$ | Basis: $\mathrm{B}_{\mathrm{pa}} \sim \mathrm{Blim}^{*} 1.4$ | Changed in ACFM 2007 (from 3800 to 3100 t ). |
| $\mathrm{F}_{\text {lim }}=0.4$ | Basis: $\mathrm{Flim}=\mathrm{F}_{\text {loss }}$ | Although poorly defined, based that there is evidence that fishing mortality in excess of 0.4 has led to a general stock decline and is only sustainable during periods of above-average recruitment. |
| $\mathrm{F}_{\mathrm{pa}}=0.3$ | Basis: $\mathrm{F}_{\mathrm{pa}}$ be set at 0.30 . | This F is considered to have a high probability of avoiding Flim. |

## Yield-per-Recruit analysis

A yield-per-recruit analysis was carried out (Table 6.6.18 and Figure 6.8.6). Current fishing mortality (0.33) is well above $\mathrm{F}_{0.1}$ ( 0.14 ). F $\mathrm{F}_{\max }$ was estimated at 0.68 , but was considered to be not well defined given flat yield-per-recruit curve. Taking the results of the stochastic yield-per-recruit analysis into account, which shows high CVs on F0.1 and very high CVs on $\mathrm{F}_{\max }$ estimates, the WG concluded that these deterministic values should not be used as a basis of management advice.

### 6.8.7 Management plans

No management plan is currently in place for Irish Sea sole.

### 6.8.8 Uncertainties and bias in assessment and forecast

## Sampling

The major fleets fishing for Irish Sea sole are sampled. Sampling is considered to be at a reasonable level. Under the DCF there is an initiative to co-ordinate sampling across the three countries involved in the fishery. One of the problems in this assessment may well be the quality of historical catch-at-age data.

## Landings

There is no reliable information on the accuracy of the landing statistics. Nevertheless, the total TAC uptake over the last 3 years was only in the range of $50-60 \%$. In this context, misreporting is not considered to be a major problem for these years.

## Discards

The absence of discard data is unlikely to affect the quality of the assessment as information from 2003, 2004 and 2005 and 2007 indicates that discarding ranges by weight vary between 0 and $8 \%$. In 2006 high discard rates were estimated for the UK beam trawl fleet, but this estimate was heavily influenced by one observation made in the fourth quarter.

## Effort

There are no indications of Irish Sea sole fisheries misreporting effort. Effort in sole targeting beam trawl fisheries has declined substantially in the last few years.

## Surveys

The UK (E\&W) September beam trawl survey appears to track year-class strength well. As previously investigated, this tuning fleet is also quite consistent in estimating year-class strength of the same year class at different ages. Therefore the Working Group had confidence in using the UK (E\&W) September survey. The bias problem in the assessment may be the result of the precise survey and less precise catch-at-age data.

## Model formulation

At the moment XSA is used to assess Irish Sea sole. In the WG of 2007 the model settings were changed which did have a considerable impact on the estimates of SSB and fishing mortality. Due to the major revisions, ACFM changed the biomass reference points in its meeting of 2007. In the last two year's update assessment (20082009) no major changes were apparent.

### 6.8.9 Recommendations for next Benchmark

Last year WGCSE recommended this stock for benchmark in 2011. This year WGCSE reiterates this recommendation. Little progress has been made thus far but the WG considers that a benchmark in 2011 is still warranted and possible.


### 6.8.1 0 Management considerations

SSB in 2009 is estimated at its lowest observed value, and well below Blim. Recruit-ment-at-age 2 has been well below average since 2001, and is estimated to remain low
in 2006-2009. The model indicates that fishing mortality has come down over the last couple of years (as did effort for most fleets fishing on Irish Sea sole), and is now close to $\mathrm{F}_{\mathrm{pa}}$.

It is not possible for the stock to reach $\mathbf{B}_{\mathrm{pa}}$ in one year. A management plan for effort reduction that can be phased in over a number of years and implemented in conjunction with technical conservation measures should be considered.

Given the successive recent low recruitment, predictions become more dependent on the assumed incoming recruitment. $55 \%$ of the predicted SSB in 2012 is based on that assumption. A GM recruitment was used, which might be an optimistic assumption given the consecutive low recruitments in recent years.

Sole is caught in a mixed fishery with other flatfish as well as gadoids. Information from observer trips indicates that discarding of sole is relatively low.

Table 6.8.1. Sole in VIIa. Nominal landings (tonnes) as officially reported by ICES, and Working Group estimates of the landings.

| Country | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 930 | 987 | 915 | 1010 | 786 | 371 | 531 | 495 | 706 | 675 | 533 | 570 | 525 | 469 | 493 | 674 | 817 | 687 | 527 | 662 | 419.3 | 305 | 216 | 257.2 |
| France | 17 | 5 | 11 | 5 | 2 | 3 | 11 | 8 | 7 | 5 | 5 | 3 | 3 | 0.5 | 3 | 4 | 4 | 4 | 1 | 3 | 1 | 1.0 | 1.0 | n/a |
| Ireland | 235 | 312 | 366 | 155 | 170 | 198 | 164 | 98 | 226 | 176 | 133 | 130 | 134 | 120 | 135 | 135 | 96 | 103 | 77 | 85 | 85 | 115 | 66.0 | 47.1 |
| Netherlands | - | - | - | - | - | - | - | - | - | - | 149 | 123 | 60 | 46 | 60 | - | - | - | - | - | - | - | - | - |
| UK (Engl.\& Wales) ${ }^{1}$ | 637 | 599 | 507 | 613 | 569 | 581 | 477 | 338 | 409 | 424 | 194 | 189 | 161 | 165 | 133 | 195 | 165 | 217 | 106 | 103 | 69 | 66 | 37 | 19.7 |
| UK (Isle of Man) | 1 | 3 | 1 | 2 | 10 | 44 | 14 | 4 | 5 | 12 | 4 | 5 | 3 | 1 | 1 | + | + | + | + | + | + | $<0.5$ | n/a |  |
| UK (N. Ireland) ${ }^{1}$ | 50 | 72 | 47 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 46 | 63 | 38 | 38 | 39 | 26 | 37 | 28 | 14 | 8 | 5 | 7 | 9 | 8 | 8 | 4 | 3 | 3 | 1 | 1 | 2 | 4 | n/a | n/a |
| Total | 1916 | 2041 | 1885 | 1823 | 1576 | 1223 | 1234 | 971 | 1367 | 1300 | 490 | 1027 | 895 | 810 | 833 | 1012 | 1085 | 1014 | 712 | 854 | 576 | 491 | 320 | 324 |
| Used by WG | 1995 | 2808 | 1999 | 1833 | 1583 | 1212 | 1259 | 1023 | 1374 | 1266 | 1002 | 1003 | 911 | 863 | 818 | 1053 | 1087 | 1014 | 699 | 855 | 569 | 492 | 333 | 324 |
| Unallocated | 79 | 767 | 114 | 10 | 7 | -11 | 25 | 52 | 7 | -34 | 512 | -24 | 16 | 54 | -15 | 41 | 2 | 0 | -13 | 1 | -7 | 1 | 13 | 0 |

Table 6.8.2. Sole in VIIa. Annual lenght distributions by fleet (2009).

| Length (cm) | UK (England \& Wales) | Belgium | Ireland |
| :---: | :---: | :---: | :---: |
|  | All gears | All gears | All gears |
| 20 |  |  |  |
| 21 | 27 |  | 47 |
| 22 | 160 | 6837 | 513 |
| 23 | 489 | 97767 | 1680 |
| 24 | 1600 | 177268 | 2054 |
| 25 | 3293 | 172129 | 2567 |
| 26 | 4332 | 141190 | 5462 |
| 27 | 3743 | 125816 | 8496 |
| 28 | 3530 | 115845 | 9056 |
| 29 | 3607 | 59004 | 8963 |
| 30 | 1667 | 57273 | 11623 |
| 31 | 2543 | 47518 | 10830 |
| 32 | 1690 | 43129 | 9616 |
| 33 | 932 | 22237 | 9896 |
| 34 | 1277 | 18269 | 8963 |
| 35 | 751 | 14750 | 7515 |
| 36 | 547 | 12977 | 7609 |
| 37 | 214 | 7134 | 4855 |
| 38 | 378 | 4255 | 5415 |
| 39 | 298 | 2692 | 3641 |
| 40 | 74 | 2177 | 3081 |
| 41 | 27 | 1333 | 1727 |
| 42 | 145 | 658 | 1447 |
| 43 | 9 | 320 | 794 |
| 44 |  | 454 | 513 |
| 45 |  | 294 | 467 |
| 46 |  | 134 | 187 |
| 47 |  | 89 | 0 |
| 48 |  |  | 140 |
| 49 |  |  |  |
| 50 |  |  |  |
| 51 |  |  |  |
| 52 |  |  |  |
| 53 |  |  |  |
| 54 |  |  |  |
| 55 |  |  |  |
| 56 |  |  |  |
| 57 |  |  |  |
| 58 |  |  |  |
| 59 |  |  |  |
| 60 |  |  |  |
| Total | 31333 | 1131549 | 127156 |

Table 6.8.3. Sole in VIIa. Catch numbers-at-age (in thousands).

|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 29 | 113 | 31 | 368 | 25 | 262 | 29 | 221 | 65 | 108 |
| 3 | 895 | 434 | 673 | 363 | 891 | 733 | 375 | 416 | 958 | 1027 |
| 4 | 1009 | 2097 | 730 | 2195 | 576 | 2386 | 1332 | 1292 | 649 | 3433 |
| 5 | 467 | 1130 | 1537 | 557 | 1713 | 539 | 2330 | 774 | 1009 | 829 |
| 6 | 1457 | 232 | 537 | 815 | 383 | 842 | 247 | 1066 | 442 | 637 |
| 7 | 289 | 878 | 172 | 267 | 422 | 157 | 544 | 150 | 638 | 326 |
| +gp | 2537 | 1887 | 1500 | 1143 | 971 | 1006 | 739 | 648 | 587 | 620 |
| TOTALNUM | 6683 | 6771 | 5180 | 5708 | 4981 | 5925 | 5596 | 4567 | 4348 | 6980 |
| TONSLAND | 1785 | 1882 | 1450 | 1428 | 1307 | 1441 | 1463 | 1147 | 1106 | 1614 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 2 | 187 | 70 | 8 | 37 | 651 | 154 | 141 | 189 | 32 | 179 |
| 3 | 939 | 580 | 346 | 165 | 786 | 1601 | 3336 | 3348 | 444 | 771 |
| 4 | 1968 | 1668 | 1241 | 998 | 380 | 1086 | 3467 | 4105 | 4752 | 775 |
| 5 | 3055 | 1480 | 1298 | 758 | 610 | 343 | 961 | 3185 | 2102 | 3978 |
| 6 | 521 | 1640 | 711 | 757 | 343 | 334 | 235 | 844 | 1310 | 1178 |
| 7 | 512 | 114 | 641 | 416 | 424 | 164 | 277 | 307 | 203 | 552 |
| +gp | 1145 | 865 | 397 | 709 | 557 | 739 | 848 | 808 | 516 | 255 |
| TOTALNUM | 8327 | 6417 | 4642 | 3840 | 3751 | 4421 | 9265 | 12786 | 9359 | 7688 |
| TONSLAND | 1941 | 1667 | 1338 | 1169 | 1058 | 1146 | 1995 | 2808 | 1999 | 1833 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 2 | 564 | 1317 | 363 | 83 | 122 | 132 | 60 | 789 | 167 | 301 |
| 3 | 1185 | 1270 | 2433 | 543 | 1342 | 920 | 469 | 713 | 1728 | 1069 |
| 4 | 986 | 841 | 918 | 1966 | 1069 | 1444 | 1188 | 474 | 466 | 1258 |
| 5 | 598 | 300 | 556 | 559 | 1578 | 737 | 741 | 710 | 256 | 297 |
| 6 | 2319 | 226 | 190 | 251 | 394 | 1010 | 430 | 408 | 315 | 115 |
| 7 | 592 | 1173 | 156 | 199 | 133 | 179 | 509 | 258 | 191 | 136 |
| +gp | 466 | 459 | 929 | 686 | 524 | 350 | 347 | 531 | 423 | 232 |
| TOTALNUM | 6710 | 5586 | 5545 | 4287 | 5162 | 4772 | 3744 | 3883 | 3546 | 3408 |
| TONSLAND | 1583 | 1212 | 1259 | 1023 | 1374 | 1266 | 1002 | 1003 | 911 | 863 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 88 | 267 | 88 | 329 | 146 | 518 | 115 | 179 | 100 | 136 |
| 3 | 1013 | 1259 | 442 | 1082 | 946 | 1066 | 630 | 462 | 391 | 513 |
| 4 | 1180 | 909 | 1329 | 1042 | 352 | 617 | 554 | 399 | 255 | 335 |
| 5 | 556 | 604 | 1122 | 704 | 332 | 408 | 233 | 242 | 204 | 129 |
| 6 | 190 | 471 | 551 | 308 | 292 | 257 | 126 | 87 | 113 | 126 |
| 7 | 66 | 68 | 194 | 155 | 91 | 167 | 142 | 37 | 65 | 71 |
| +gp | 224 | 238 | 119 | 201 | 78 | 248 | 250 | 212 | 114 | 58 |
| TOTALNUM | 3317 | 3816 | 3845 | 3821 | 2237 | 3281 | 2050 | 1618 | 1242 | 1368 |
| TONSLAND | 818 | 1053 | 1090 | 1014 | 709 | 855 | 569 | 492 | 332 | 324 |
| SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table 6.8.4. Sole in VIIa. Catch weights-at-age (kg).

|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.130 | 0.152 | 0.126 | 0.151 | 0.138 | 0.130 | 0.120 | 0.085 | 0.093 | 0.134 |
| 3 | 0.153 | 0.178 | 0.164 | 0.178 | 0.174 | 0.172 | 0.161 | 0.146 | 0.147 | 0.165 |
| 4 | 0.178 | 0.204 | 0.201 | 0.204 | 0.209 | 0.210 | 0.200 | 0.202 | 0.197 | 0.199 |
| 5 | 0.204 | 0.230 | 0.237 | 0.230 | 0.241 | 0.244 | 0.239 | 0.251 | 0.243 | 0.234 |
| 6 | 0.232 | 0.257 | 0.272 | 0.256 | 0.272 | 0.275 | 0.276 | 0.293 | 0.286 | 0.271 |
| 7 | 0.260 | 0.284 | 0.306 | 0.283 | 0.301 | 0.303 | 0.313 | 0.330 | 0.326 | 0.311 |
| +gp | 0.377 | 0.419 | 0.417 | 0.392 | 0.396 | 0.367 | 0.457 | 0.387 | 0.429 | 0.451 |
| SOPCOFAC | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 2 | 0.146 | 0.162 | 0.112 | 0.189 | 0.191 | 0.144 | 0.122 | 0.135 | 0.111 | 0.125 |
| 3 | 0.169 | 0.183 | 0.171 | 0.212 | 0.225 | 0.189 | 0.164 | 0.164 | 0.147 | 0.163 |
| 4 | 0.193 | 0.207 | 0.225 | 0.238 | 0.257 | 0.231 | 0.203 | 0.196 | 0.183 | 0.201 |
| 5 | 0.219 | 0.234 | 0.275 | 0.266 | 0.288 | 0.272 | 0.241 | 0.231 | 0.218 | 0.237 |
| 6 | 0.247 | 0.264 | 0.321 | 0.298 | 0.318 | 0.310 | 0.277 | 0.268 | 0.252 | 0.271 |
| 7 | 0.275 | 0.296 | 0.362 | 0.332 | 0.347 | 0.346 | 0.311 | 0.308 | 0.286 | 0.304 |
| +gp | 0.380 | 0.452 | 0.456 | 0.458 | 0.409 | 0.430 | 0.407 | 0.462 | 0.419 | 0.389 |
| SOPCOFAC | 1.001 | 1.000 | 1.000 | 1.000 | 1.000 | 0.999 | 0.999 | 1.000 | 0.999 | 1.000 |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 2 | 0.135 | 0.133 | 0.149 | 0.102 | 0.175 | 0.129 | 0.156 | 0.154 | 0.187 | 0.179 |
| 3 | 0.162 | 0.172 | 0.177 | 0.156 | 0.198 | 0.182 | 0.193 | 0.197 | 0.209 | 0.217 |
| 4 | 0.192 | 0.208 | 0.207 | 0.205 | 0.227 | 0.232 | 0.228 | 0.237 | 0.234 | 0.252 |
| 5 | 0.227 | 0.241 | 0.239 | 0.248 | 0.261 | 0.277 | 0.263 | 0.275 | 0.263 | 0.285 |
| 6 | 0.265 | 0.272 | 0.274 | 0.285 | 0.301 | 0.318 | 0.296 | 0.311 | 0.295 | 0.314 |
| 7 | 0.307 | 0.300 | 0.310 | 0.318 | 0.346 | 0.356 | 0.327 | 0.345 | 0.331 | 0.341 |
| +gp | 0.414 | 0.345 | 0.379 | 0.370 | 0.509 | 0.451 | 0.410 | 0.407 | 0.440 | 0.399 |
| SOPCOFAC | 1.000 | 1.000 | 0.999 | 0.999 | 1.001 | 1.000 | 1.000 | 1.002 | 1.000 | 1.001 |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.143 | 0.184 | 0.163 | 0.143 | 0.188 | 0.203 | 0.209 | 0.219 | 0.197 | 0.154 |
| 3 | 0.190 | 0.231 | 0.212 | 0.206 | 0.257 | 0.231 | 0.234 | 0.255 | 0.228 | 0.217 |
| 4 | 0.235 | 0.273 | 0.257 | 0.262 | 0.318 | 0.258 | 0.259 | 0.289 | 0.257 | 0.244 |
| 5 | 0.276 | 0.308 | 0.298 | 0.310 | 0.372 | 0.284 | 0.284 | 0.321 | 0.284 | 0.299 |
| 6 | 0.315 | 0.338 | 0.334 | 0.352 | 0.418 | 0.308 | 0.309 | 0.352 | 0.309 | 0.264 |
| 7 | 0.351 | 0.362 | 0.367 | 0.386 | 0.456 | 0.331 | 0.334 | 0.382 | 0.332 | 0.269 |
| +gp | 0.443 | 0.393 | 0.423 | 0.420 | 0.505 | 0.374 | 0.399 | 0.460 | 0.383 | 0.3319 |
| SOPCOFAC | 1.000 | 1.001 | 1.000 | 1.000 | 1.000 | 1.002 | 1.003 | 1.000 | 0.999 | 0.9994 |

Table 6.8.5. Sole in VIIa. Stock weights-at-age (kg).

|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.118 | 0.139 | 0.106 | 0.138 | 0.119 | 0.108 | 0.100 | 0.052 | 0.065 | 0.119 |
| 3 | 0.141 | 0.165 | 0.145 | 0.164 | 0.156 | 0.151 | 0.141 | 0.116 | 0.120 | 0.149 |
| 4 | 0.166 | 0.191 | 0.183 | 0.191 | 0.192 | 0.191 | 0.181 | 0.175 | 0.172 | 0.182 |
| 5 | 0.191 | 0.217 | 0.219 | 0.217 | 0.225 | 0.228 | 0.220 | 0.227 | 0.220 | 0.216 |
| 6 | 0.218 | 0.244 | 0.255 | 0.243 | 0.257 | 0.260 | 0.258 | 0.273 | 0.265 | 0.252 |
| 7 | 0.246 | 0.271 | 0.289 | 0.270 | 0.287 | 0.290 | 0.295 | 0.312 | 0.306 | 0.291 |
| +gp | 0.360 | 0.405 | 0.403 | 0.379 | 0.385 | 0.361 | 0.442 | 0.3815 | 0.417 | 0.428 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 2 | 0.135 | 0.152 | 0.081 | 0.179 | 0.174 | 0.121 | 0.101 | 0.121 | 0.093 | 0.105 |
| 3 | 0.157 | 0.172 | 0.142 | 0.200 | 0.208 | 0.167 | 0.143 | 0.149 | 0.129 | 0.144 |
| 4 | 0.181 | 0.195 | 0.198 | 0.224 | 0.241 | 0.210 | 0.183 | 0.180 | 0.165 | 0.182 |
| 5 | 0.206 | 0.220 | 0.251 | 0.252 | 0.273 | 0.252 | 0.222 | 0.213 | 0.200 | 0.219 |
| 6 | 0.233 | 0.249 | 0.299 | 0.282 | 0.303 | 0.291 | 0.259 | 0.249 | 0.235 | 0.254 |
| 7 | 0.261 | 0.280 | 0.342 | 0.315 | 0.332 | 0.328 | 0.294 | 0.287 | 0.269 | 0.288 |
| +gp | 0.363 | 0.430 | 0.443 | 0.436 | 0.396 | 0.415 | 0.393 | 0.437 | 0.403 | 0.374 |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 2 | 0.123 | 0.113 | 0.135 | 0.073 | 0.165 | 0.101 | 0.136 | 0.132 | 0.177 | 0.159 |
| 3 | 0.148 | 0.153 | 0.162 | 0.130 | 0.186 | 0.156 | 0.174 | 0.176 | 0.198 | 0.199 |
| 4 | 0.176 | 0.190 | 0.192 | 0.181 | 0.212 | 0.207 | 0.211 | 0.217 | 0.221 | 0.235 |
| 5 | 0.209 | 0.225 | 0.223 | 0.227 | 0.243 | 0.255 | 0.246 | 0.257 | 0.248 | 0.269 |
| 6 | 0.245 | 0.257 | 0.256 | 0.267 | 0.280 | 0.298 | 0.279 | 0.294 | 0.279 | 0.300 |
| 7 | 0.286 | 0.286 | 0.292 | 0.302 | 0.323 | 0.338 | 0.312 | 0.328 | 0.312 | 0.328 |
| +gp | 0.388 | 0.334 | 0.359 | 0.362 | 0.478 | 0.440 | 0.397 | 0.393 | 0.418 | 0.391 |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.119 | 0.158 | 0.137 | 0.109 | 0.151 | 0.189 | 0.196 | 0.201 | 0.181 | 0.159 |
| 3 | 0.167 | 0.208 | 0.188 | 0.175 | 0.224 | 0.218 | 0.221 | 0.237 | 0.213 | 0.199 |
| 4 | 0.213 | 0.253 | 0.235 | 0.235 | 0.289 | 0.245 | 0.246 | 0.272 | 0.243 | 0.235 |
| 5 | 0.256 | 0.291 | 0.278 | 0.287 | 0.346 | 0.271 | 0.271 | 0.305 | 0.271 | 0.269 |
| 6 | 0.296 | 0.324 | 0.317 | 0.332 | 0.396 | 0.296 | 0.296 | 0.337 | 0.297 | 0.300 |
| 7 | 0.334 | 0.351 | 0.351 | 0.370 | 0.438 | 0.320 | 0.321 | 0.367 | 0.321 | 0.328 |
| +gp | 0.430 | 0.386 | 0.413 | 0.415 | 0.497 | 0.365 | 0.386 | 0.447 | 0.375 | 0.391 |

Table 6.8.6. Sole in VIIa. Tuning-series (values in bold are used in the assessment).

| BEL-BEAM | Belgium Beam trawl (Effort = Corrected formula) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 2005 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |
| 4 | 14 |  |  |  |  |  |  |  |  |  |  |
| 12.3 | 1045 | 275 | 393 | 69 | 105 | 94 | 61 | 72 | 11 | 15 | 64 |
| 11.8 | 568 | 1066 | 80 | 263 | 64 | 58 | 35 | 5 | 56 | 5 | 5 |
| 10.7 | 434 | 307 | 509 | 76 | 93 | 45 | 23 | 20 | 2 | 35 | 32 |
| 9.9 | 169 | 304 | 155 | 258 | 41 | 90 | 12 | 29 | 12 | 7 | 17 |
| 11.2 | 1455 | 510 | 323 | 193 | 162 | 37 | 36 | 9 | 41 | 0 | 0 |
| 16.7 | 958 | 1644 | 296 | 268 | 247 | 210 | 30 | 64 | 31 | 14 | 7 |
| 22.6 | 909 | 721 | 998 | 62 | 92 | 44 | 161 | 13 | 92 | 10 | 8 |
| 19.5 | 451 | 608 | 378 | 394 | 52 | 64 | 11 | 29 | 24 | 5 | 0 |
| 20.5 | 259 | 310 | 394 | 238 | 216 | 44 | 38 | 28 | 49 | 3 | 26 |
| 12 | 107 | 204 | 143 | 188 | 91 | 121 | 2 | 1 | 4 | 14 | 0 |
| 19.6 | 606 | 171 | 186 | 99 | 150 | 125 | 83 | 27 | 13 | 4 | 23 |
| 38 | 1531 | 468 | 138 | 135 | 90 | 104 | 69 | 69 | 20 | 8 | 21 |
| 43.2 | 1527 | 881 | 297 | 167 | 69 | 39 | 54 | 59 | 40 | 13 | 9 |
| 30.5 | 2027 | 1012 | 480 | 21 | 33 | 37 | 34 | 42 | 35 | 0 | 7 |
| 34 | 376 | 2423 | 751 | 250 | 59 | 15 | 9 | 2 | 14 | 0 | 1 |
| 36.1 | 307 | 223 | 1263 | 276 | 142 | 13 | 9 | 11 | 11 | 8 | 5 |
| 13.8 | 253 | 78 | 60 | 588 | 115 | 40 | 16 | 1 | 1 | 11 | 3 |
| 23.9 | 298 | 330 | 68 | 40 | 203 | 93 | 36 | 12 | 0 | 0 | 0 |
| 24.5 | 862 | 253 | 149 | 89 | 79 | 160 | 66 | 77 | 0 | 0 | 0 |
| 31 | 680 | 786 | 164 | 103 | 39 | 117 | 58 | 19 | 15 | 0 | 7 |
| 26.2 | 729 | 366 | 410 | 52 | 27 | 6 | 28 | 15 | 6 | 11 | 3 |
| 21.6 | 537 | 334 | 241 | 219 | 53 | 13 | 11 | 14 | 9 | 7 | 2 |
| 28.5 | 270 | 376 | 180 | 162 | 134 | 28 | 27 | 15 | 9 | 8 | 1 |
| 23.3 | 248 | 146 | 142 | 89 | 73 | 62 | 20 | 20 | 9 | 10 | 3 |
| 21.7 | 693 | 199 | 65 | 50 | 37 | 21 | 17 | 9 | 6 | 4 | 6 |
| 18.6 | 685 | 220 | 107 | 31 | 15 | 33 | 13 | 7 | 9 | 0.6 | 8 |
| 30.5 | 600 | 284 | 248 | 39 | 35 | 44 | 33 | 1 | 3 | 0.2 | 4 |
| 38.6 | 1138 | 814 | 349 | 109 | 30 | 9 | 2 | 1 | 1 | 1 | 0 |
| 24.45 | 724 | 436 | 196 | 84 | 20 | 7 | 2 | 1 | 0 | 2 | 1 |
| 25.58 | 313 | 197 | 159 | 47 | 12 | 11 | 6 | 3 | 0 | 0 | 0 |
| 32.15 | 505 | 342 | 156 | 71 | 87 | 9 | 7 | 1 | 13 | 2 | 1 |
| E+W September beam trawl survey |  |  |  |  |  |  |  |  |  |  |  |
| 1988 | 2009 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |  |  |
| 1 | 9 |  |  |  |  |  |  |  |  |  |  |
| 100.062 | 118 | 196 | 180 | 410 | 76 | 40 | 4 | 0 | 4 |  |  |
| 129.71 | 218 | 304 | 180 | 74 | 284 | 56 | 32 | 8 | 6 |  |  |
| 128.969 | 1712 | 534 | 122 | 42 | 88 | 194 | 40 | 20 | 6 |  |  |
| 123.78 | 148 | 1286 | 122 | 26 | 16 | 14 | 55 | 19 | 7 |  |  |
| 129.525 | 220 | 309 | 657 | 142 | 34 | 22 | 7 | 75 | 17 |  |  |
| 131.192 | 83 | 330 | 143 | 211 | 40 | 17 | 7 | 16 | 36 |  |  |
| 124.892 | 60 | 408 | 203 | 73 | 132 | 49 | 11 | 13 | 6 |  |  |
| 126.004 | 246 | 154 | 253 | 110 | 30 | 67 | 12 | 5 | 5 |  |  |
| 126.004 | 886 | 126 | 32 | 76 | 46 | 23 | 31 | 8 | 2 |  |  |
| 126.004 | 1158 | 577 | 72 | 24 | 55 | 27 | 16 | 30 | 7 |  |  |
| 126.004 | 539 | 716 | 292 | 18 | 6 | 24 | 23 | 5 | 18 |  |  |
| 126.004 | 385 | 293 | 255 | 203 | 29 | 8 | 26 | 5 | 6 |  |  |
| 126.004 | 354 | 464 | 147 | 219 | 91 | 13 | 2 | 13 | 6 |  |  |
| 126.004 | 91 | 284 | 192 | 65 | 96 | 64 | 6 | 3 | 12 |  |  |
| 126.004 | 205 | 61 | 121 | 126 | 42 | 79 | 49 | 2 | 1 |  |  |
| 126.004 | 242 | 210 | 51 | 97 | 81 | 40 | 43 | 26 | 1 |  |  |
| 126.004 | 406 | 240 | 119 | 27 | 77 | 45 | 41 | 17 | 19 |  |  |
| 122.298 | 53 | 165 | 69 | 25 | 13 | 35 | 25 | 4 | 6 |  |  |
| 126.004 | 107 | 110 | 90 | 45 | 36 | 9 | 16 | 15 | 10 |  |  |
| 126.004 | 125 | 93 | 49 | 57 | 41 | 11 | 4 | 6 | 12 |  |  |
| 122.298 | 126 | 125 | 60 | 21 | 43 | 23 | 6 | 2 | 9 |  |  |
| 126.004 | 57 | 150 | 68 | 39 | 23 | 30 | 12 | 7 | 1 |  |  |

Table 6.8.6. Sole in VIIa. Continued.

| E+W March beam trawl survey |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 1999 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.15 | 0.25 |  |  |  |  |  |  |  |  |  |  |
| 1 | 9 |  |  |  |  |  |  |  |  |  |  |  |  |
| 126.931 | 18 | 337 | 147 | 332 | 73 | 15 | 17 | 10 | 41 |  |  |  |  |
| 115.442 | 8 | 354 | 208 | 69 | 151 | 51 | 14 | 11 | 9 |  |  |  |  |
| 126.189 | 24 | 96 | 186 | 140 | 30 | 104 | 27 | 10 | 8 |  |  |  |  |
| 134.343 | 651 | 114 | 49 | 110 | 78 | 32 | 54 | 10 | 12 |  |  |  |  |
| 121.742 | 130 | 417 | 33 | 17 | 69 | 23 | 11 | 46 | 17 |  |  |  |  |
| 130.081 | 47 | 421 | 330 | 39 | 19 | 48 | 27 | 12 | 37 |  |  |  |  |
| 130.822 | 45 | 227 | 284 | 177 | 14 | 4 | 34 | 12 | 7 |  |  |  |  |
| UK(E+W) Beam trawl |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | 2008 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 25.838 | 267 | 426 | 212 | 84 | 58 | 218 | 53 | 34 | 4 | 1 | 2 | 1 | 0 |
| 23.399 | 36 | 460 | 176 | 68 | 37 | 32 | 121 | 34 | 38 | 3 | 1 | 0 | 0 |
| 21.503 | 11 | 74 | 355 | 98 | 36 | 48 | 25 | 34 | 13 | 22 | 5 | 2 | 4 |
| 20.145 | 24 | 228 | 150 | 234 | 87 | 17 | 25 | 19 | 42 | 10 | 17 | 1 | 0 |
| 20.392 | 47 | 239 | 231 | 130 | 199 | 55 | 11 | 22 | 5 | 34 | 10 | 11 | 3 |
| 13.32 | 0 | 13 | 109 | 98 | 49 | 100 | 37 | 9 | 8 | 6 | 14 | 8 | 3 |
| 10.76 | 0 | 111 | 50 | 81 | 58 | 24 | 46 | 34 | 12 | 12 | 0 | 8 | 1 |
| 10.386 | 43 | 219 | 40 | 28 | 49 | 31 | 12 | 22 | 11 | 9 | 2 | 1 | 0 |
| 11.016 | 53 | 115 | 134 | 12 | 15 | 25 | 10 | 9 | 14 | 9 | 0 | 1 | 2 |
| 6.275 | 16 | 90 | 84 | 82 | 9 | 6 | 10 | 5 | 5 | 7 | 2 | 1 | 1 |
| 12.495 | 33 | 184 | 100 | 145 | 107 | 12 | 4 | 17 | 12 | 10 | 6 | 4 | 2 |
| 8.017 | 4 | 63 | 152 | 50 | 79 | 47 | 5 | 4 | 6 | 3 | 1 | 1 | 1 |
| 13.996 | 28 | 63 | 178 | 149 | 78 | 52 | 72 | 7 | 5 | 8 | 3 | 7 | 14 |
| 7.396 | 54 | 61 | 29 | 43 | 25 | 12 | 10 | 5 | 1 | 1 | 4 | 0 | 1 |
| 11.406 | 10 | 81 | 44 | 16 | 45 | 37 | 17 | 10 | 17 | 3 | 0 | 3 | 3 |
| 4.649 | 7 | 28 | 33 | 11 | 5 | 10 | 12 | 7 | 9 | 5 | 2 | 0 | 1 |
| 3.197 | 22 | 20 | 34 | 17 | 6 | 1 | 7 | 7 | 6 | 3 | 2 | 1 | 1 |
| 1.302 | 1 | 11 | 5 | 7 | 12 | 1 | 2 | 4 | 3 | 4 | 0 | 3 | 1 |
| IR-OTB : Irish Otter trawl - Effort in hours - VIla Sole numbers at age - Year |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995 | 2005 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 10 |  |  |  |  |  |  |  |  |  |  |  |  |
| 70682 | 6.8 | 17.7 | 25.5 | 9.2 | 25.8 | 3.6 | 0.8 | 1.5 | 1.9 | 1995 |  |  |  |
| 58166 | 0 | 5.7 | 12.9 | 12.7 | 4.7 | 4.7 | 2.2 | 0.2 | 0 | 1996 |  |  |  |
| 75029 | 27.8 | 10.2 | 4.1 | 9.2 | 6.4 | 3.5 | 3.9 | 1 | 0.2 | 1997 |  |  |  |
| 81073 | 5.5 | 40.7 | 14.7 | 6.6 | 12.3 | 5.4 | 2.7 | 4.1 | 1 | 1998 |  |  |  |
| 93221 | 26.6 | 36.8 | 30.9 | 5.1 | 3.8 | 5.3 | 2.4 | 0.5 | 1.2 | 1999 |  |  |  |
| 64320 | 1.6 | 13.2 | 13.4 | 11 | 3.4 | 1.1 | 1 | 0.4 | 0 | 2000 |  |  |  |
| 77541 | 0.2 | 6.1 | 18.6 | 18.6 | 10.8 | 2.1 | 4.1 | 1.3 | 0.3 | 2001 |  |  |  |
| 39996 | 20.3 | 20 | 30.2 | 16.4 | 8.2 | 2.9 | 2.4 | 1.4 | 0.5 | 2002 |  |  |  |
| 73854 | 0.9 | 35.9 | 21.7 | 9.8 | 3.3 | 0.5 | 0.8 | 0.2 | 0.2 | 2003 |  |  |  |
| 72507 | 9 | 15.1 | 4.1 | 3.2 | 1.9 | 1.6 | 0.3 | 0.2 | 0.1 | 2004 |  |  |  |
| \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\# |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 31142 | 4 | 1.7 | 1.6 | 1.6 | 0.6 | 0.1 | 0 | 0 | 0 | 2005 |  |  |  |
| Please note the 2005 data is based only on Q3 and Q4 data and has not been raised to annual effort. It should not be included as part of this time series. |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 6.8.7. Sole in VIIa. Effort and cpue series.

| Year | CPUE |  |  |  |  |  |  | Effort |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { Belgium } \\ \\ \text { beam } \end{gathered}$ | $U K(E+W)^{3}$ <br> beam otter |  | $U K^{5}$ <br> beam survey |  | Ireland otter beam |  | $\begin{gathered} \text { Belgium } \\ \text { beam } \\ \hline \end{gathered}$ | $\mathrm{UK}(\mathrm{E}+\mathrm{W})^{4}$ |  | Ireland ${ }^{6}$ |  |
|  | Whole year | Whole year | Whole year | Sept | March | Whole year | Whole year | Whole year | Whole year | Whole year | Whole Year | Whole Year |
| 1972 | - | - | 1.06 | - | - | - | - | - | - | 128.4 | - | - |
| 1973 | - | - | 1.06 | - | - | - | - | - | - | 147.6 | - | - |
| 1974 | - | - | 1.09 | - | - | - | - | - | - | 115.2 | - | - |
| 1975 | 21.4 | - | 1.39 | - | - | - | - | 28.4 | - | 130.7 | - | - |
| 1976 | 23.1 | - | 0.94 | - | - | - | - | 24.9 | - | 122.3 | - | - |
| 1977 | 19.8 | - | 0.80 | - | - | - | - | 22.1 | - | 101.9 | - | - |
| 1978 | 18.1 | 34.32 | 1.04 | - | - | - | - | 17.5 | 0.9 | 89.1 | - | - |
| 1979 | 33.4 | 32.01 | 1.43 | - | - | - | - | 20.4 | 1.7 | 89.9 | - | - |
| 1980 | 28.2 | 31.70 | 1.01 | - | - | - | - | 32.0 | 4.3 | 107.0 | - | - |
| 1981 | 22.2 | 21.32 | 0.75 | - | - | - | - | 36.5 | 6.4 | 107.1 | - | - |
| 1982 | 22.0 | 29.94 | 0.53 | - | - | - | - | 26.5 | 5.5 | 127.2 | - | - |
| 1983 | 13.9 | 37.31 | 0.57 | - | - | - | - | 28.7 | 2.8 | 88.1 | - | - |
| 1984 | 22.5 | 16.24 | 0.71 | - | - | - | - | 17.5 | 4.1 | 103.1 | - | - |
| 1985 | 20.6 | 17.34 | 0.56 | - | - | - | - | 27.0 | 7.4 | 102.9 | - | - |
| 1986 | 19.1 | 19.23 | 0.84 | - | - | - | - | 44.5 | 17.0 | 90.3 | - | - |
| 1987 | 17.7 | 14.82 | 0.77 | - | - | - | - | 51.6 | 22.0 | 130.6 | - | - |
| 1988 | 21.3 | 11.81 | 0.46 | 158.7 | - | - | - | 38.2 | 18.6 | 132.0 | - | - |
| 1989 | 21.9 | 9.17 | 0.70 | 145.9 | - | - | - | 42.2 | 25.3 | 139.5 | - | - |
| 1990 | 17.5 | 9.52 | 0.61 | 190.1 | - | - | - | 42.4 | 31.0 | 117.1 | - | - |
| 1991 | 18.7 | 10.43 | 1.12 | 170.5 | - | - | - | 17.1 | 25.8 | 107.3 | - | - |
| 1992 | 19.2 | 9.50 | 1.02 | 158.3 | - | - | - | 25.1 | 23.4 | 96.8 | - | - |
| 1993 | 20.0 | 7.60 | 0.54 | 97.3 | 104.7 | - | - | 23.9 | 21.5 | 78.9 | - | - |
| 1994 | 19.1 | 11.76 | 0.74 | 107.7 | 91.9 |  | - | 32.5 | 20.1 | 43.0 | - | - |
| 1995 | 18.1 | 14.96 | 0.95 | 89.5 | 79.3 | 0.38 | 12.69 | 28.6 | 20.9 | 43.1 | 80.3 | 8.64 |
| 1996 | 17.7 | 9.44 | 0.53 | 86.8 | - | 0.25 | 14.94 | 23.2 | 13.3 | 42.2 | 64.8 | 6.26 |
| 1997 | 16.6 | 10.49 | 0.73 | 151.2 | 63.3 | 0.23 | 8.53 | 30.7 | 10.8 | 39.9 | 92.2 | 9.86 |
| 1998 | 19.0 | 8.42 | 0.48 | 140.8 | 89.3 | 0.38 | 7.77 | 24.7 | 10.4 | 36.9 | 93.5 | 11.58 |
| 1999 | 19.5 | 9.94 | 0.60 | 107.3 | - | 0.29 | 9.22 | 22.7 | 11.0 | 22.9 | 110.3 | 14.67 |
| 2000 | 15.5 | 12.90 | 0.44 | 122.6 | - | 0.29 | 8.49 | 26.0 | 6.3 | 27.0 | 82.7 | 11.42 |
| 2001 | 15.0 | 11.72 | 0.15 | 96.9 | - | 0.38 | 7.86 | 36.8 | 12.5 | 32.8 | 77.5 | 13.13 |
| 2002 | 15.0 | 16.73 | 1.48 | 76.0 | - | 0.32 | 4.67 | 47.0 | 8.0 | 24.8 | 77.9 | 17.67 |
| 2003 | 14.8 | 13.20 | 0.15 | 88.6 | - | 0.34 | 4.20 | 43.6 | 14.0 | 23.9 | 73.9 | 18.70 |
| 2004 | 15.4 | 13.86 | 0.17 | 98.9 | - | 0.14 | 4.31 | 32.0 | 7.4 | 23.5 | 72.5 | 14.19 |
| 2005 | 16.7 | 9.14 | 0.19 | 48.9 | - | 0.16 | 4.70 | 37.5 | 11.4 | 16.7 | 68.3 | 14.67 |
| 2006 | 15.7 | 7.83 | 0.52 | 52.6 | - | 0.16 | 6.00 | 24.6 | 4.6 | 5.2 | 66.2 | 12.20 |
| 2007 | 13.7 | 16.38 | 0.42 | 53.0 | - | 0.37 | 6.39 | 19.4 | 3.2 | 4.4 | 73.1 | 14.00 |
| 2008 | 19.5 | 15.25 | 0.30 | 50.7 | - | 0.20 | 6.13 | 9.6 | 1.3 | 2.7 | 58.8 | 9.46 |
| 2009* | 23.2 | 18.88 | 0.22 | 45.8 | - | 0.28 | 4.53 | 11.1 | 0.5 | 1.5 | 41.5 | 7.59 |
| All CPUE values in $\mathrm{Kg} / \mathrm{hr}$ except UK beam survey ( $\mathrm{Kg} / 100 \mathrm{~km}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{1} \mathrm{Kg} / 000 \mathrm{hr}$ |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2} 000$ hours fishing |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3} \mathrm{Kg} / 000 \mathrm{hr}$ fished (GRT corrected $>40$ vessels) |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4} 000$ 'hours fished (GRT corrected $>40$ ' vessels) |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{5} \mathrm{Kg} / 100 \mathrm{~km}$ fished |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{6} 000$ 'hours |  |  |  |  |  |  |  |  |  |  |  |  |
| * Provisional |  |  |  |  |  |  |  |  |  |  |  |  |

## Table 6.8.8. Sole in VIIa. Diagnostics.

| Lowestoft VPA Version 3.1 |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 28/05/2010 14:06 |  |  |  |  |  |  |  |  |  |  |  |
| Extended Survivors Analysis |  |  |  |  |  |  |  |  |  |  |  |
| IRISH SEA SOLE 2010 WG COMBSEX PLUSGROUP. |  |  |  |  |  |  |  |  |  |  |  |
| CPUE data from file SOL7ATUN.DAT |  |  |  |  |  |  |  |  |  |  |  |
| Catch data for 40 years. 1970 to 2009. Ages 2 to 8 . |  |  |  |  |  |  |  |  |  |  |  |
| Fleet |  | Firs <br> year |  |  |  |  |  |  |  |  |  |
| E+W September beam t |  | 1988 | 2009 | 2 | 7 | 0.75 | 0.85 |  |  |  |  |
| E+W March beam trawl |  | 1993 | 2009 | 2 | 7 | 0.15 | 0.25 |  |  |  |  |
| Time series weights : |  |  |  |  |  |  |  |  |  |  |  |
| Tapered time weighting applied |  |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : |  |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of stock size for all ages |  |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of age for ages $>=7$ |  |  |  |  |  |  |  |  |  |  |  |
| Terminal population estimation : |  |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk towards the mean $F$ of the final 5 years or the 3 oldest ages. |  |  |  |  |  |  |  |  |  |  |  |
| S.E. of the mean to which the estimates are shrunk $=1.500$ |  |  |  |  |  |  |  |  |  |  |  |
| Minimum standard error for population estimates derived from each fleet $=.300$ |  |  |  |  |  |  |  |  |  |  |  |
| Prior weighting not applied |  |  |  |  |  |  |  |  |  |  |  |
| Tuning had not converged after 30 iterations |  |  |  |  |  |  |  |  |  |  |  |
| Total absolute residual between iterations |  |  |  |  |  |  |  |  |  |  |  |
| 29 and $30=.00953$ |  |  |  |  |  |  |  |  |  |  |  |
| Final year $F$ values |  |  |  |  |  |  |  |  |  |  |  |
| Age |  | 2 | 3 | 4 | 5 | 6 | 7 |  |  |  |  |
| Iteration 29 |  | 0.0616 | 0.3438 | 0.383 | 0.3057 | 0.2034 | 0.2379 |  |  |  |  |
| Iteration 30 |  | 0.0615 | 0.3429 | 0.3814 | 0.3039 | 0.2016 | 0.2347 |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 0.55 | 0.6 | 0.65 | 0.7 | 0.75 | 0.8 | 0.85 | 0.9 | 0.95 | 1 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |  |
| Age |  | 2000.000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 2 | 0.016 | 0.075 | 0.034 | 0.121 | 0.049 | 0.189 | 0.075 | 0.099 | 0.050 | 0.061 |
|  | 3 | 0.237 | 0.29 | 0.154 | 0.633 | 0.525 | 0.519 | 0.328 | 0.421 | 0.289 | 0.343 |
|  | 4 | 0.342 | 0.309 | 0.498 | 0.568 | 0.383 | 0.688 | 0.496 | 0.317 | 0.384 | 0.381 |
|  | 5 | 0.253 | 0.262 | 0.68 | 0.475 | 0.314 | 0.909 | 0.532 | 0.371 | 0.237 | 0.304 |

Table 6.8.8.cont. Sole in VIIa. Diagnostics.


Table 6.8.8.cont. Sole in VIIa. Diagnostics.


Table 6.8.8.cont. Sole in VIIa. Diagnostics.

| Year class $=2007$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| E+W September beam t | 2099 | 0.361 | 0 | 0 | 1 | 0.942 | 0.06 |
| E+W March beam trawl | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 1337 | 1.5 |  |  |  | 0.058 | 0.092 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Sunvivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 2044 | 0.35 | 0.11 | 2 | 0.31 | 0.061 |  |  |
| Age 3 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |
| Year class $=2006$ |  |  |  |  |  |  |  |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| E+W September beam t | 1207 | 0.233 | 0.031 | 0.13 | 2 | 0.966 | 0.34 |
| E+W March beam trawl | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 941 | 1.5 |  |  |  | 0.034 | 0.418 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 1197 | 0.23 | 0.04 | 3 | 0.168 | 0.343 |  |  |
| 1 |  |  |  |  |  |  |  |
| Age 4 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |
| Year class $=2005$ |  |  |  |  |  |  |  |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| E+W September beam t | 695 | 0.202 | 0.105 | 0.52 | 3 | 0.969 | 0.378 |
| E+W March beam trawl | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 553 | 1.5 |  |  |  | 0.031 | 0.455 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Sunivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 690 | 0.2 | 0.09 | 4 | 0.434 | 0.381 |  |  |
| Age 5 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |
| Year class $=2004$ |  |  |  |  |  |  |  |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| E+W September beam t | 354 | 0.196 | 0.068 | 0.35 | 4 | 0.967 | 0.297 |
| E+W March beam trawl | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 202 | 1.5 |  |  |  | 0.033 | 0.474 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors at end of year | Int s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ |  | Var <br> Ratio | F |  |  |
|  | 0.2 | 0.08 | 5 | 0.393 | 0.304 |  |  |

Table 6.8.8.cont. Sole in VIIa. Diagnostics.

| Age 6 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year class $=2003$ |  |  |  |  |  |  |  |  |
| Fleet |  | Int | Ext | Var | N |  | Scaled | Estimated |
|  |  | s.e | s.e | Ratio |  |  | Weights | F |
| E+W September beam t | 551 | 0.173 | 0.035 | 0.2 |  | 5 | 0.979 | 0.197 |
| E+W March beam trawl | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| F shrinkage mean | 241 | 1.5 |  |  |  |  | 0.021 | 0.404 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | nt | Ext | N | Var | F |  |  |  |
| at end of year |  | s.e |  | Ratio |  |  |  |  |
| 542 | 0.17 | 0.06 | 6 | 0.362 |  | 0.202 |  |  |
| Age 7 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| Year class $=2002$ |  |  |  |  |  |  |  |  |
| Fleet |  | Int | Ext | Var Ratio | N |  | Scaled <br> Weights | $\begin{aligned} & \text { Estimated } \\ & \text { F } \end{aligned}$ |
|  |  | s.e | s.e |  |  |  |  |  |
| E+W September beam t | 261 | 0.167 | 0.07 | 0.42 |  | 6 | 0.979 | 0.23 |
| E+W March beam trawl | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| F shrinkage mean | 196 | 1.5 |  |  |  |  | 0.021 | 0.296 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors at end of year | nt | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | N | Var <br> Ratio | F |  |  |  |
|  | 0.17 | 0.07 | 7 | 0.393 |  | 0.235 |  |  |
|  |  |  |  |  |  |  |  |  |

Table 6.8.9. Sole in VIIa. Fishing mortality.

|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.0083 | 0.0117 | 0.0103 | 0.0299 | 0.0045 | 0.042 | 0.0079 | 0.0148 | 0.0075 | 0.0128 |  |
| 3 | 0.1196 | 0.148 | 0.0809 | 0.1435 | 0.0846 | 0.1574 | 0.0703 | 0.1348 | 0.0741 | 0.1419 |  |
| 4 | 0.2956 | 0.3988 | 0.3518 | 0.362 | 0.3156 | 0.303 | 0.4189 | 0.325 | 0.2862 | 0.3632 |  |
| 5 | 0.4444 | 0.5544 | 0.5057 | 0.4393 | 0.472 | 0.4841 | 0.4812 | 0.4066 | 0.4027 | 0.6307 |  |
| 6 | 0.4292 | 0.3671 | 0.4929 | 0.4872 | 0.5433 | 0.397 | 0.3789 | 0.3747 | 0.3808 | 0.4246 |  |
| 7 | 0.3909 | 0.4415 | 0.4516 | 0.4309 | 0.4451 | 0.3959 | 0.4277 | 0.3699 | 0.3576 | 0.4744 |  |
| +gp | 0.3909 | 0.4415 | 0.4516 | 0.4309 | 0.4451 | 0.3959 | 0.4277 | 0.3699 | 0.3576 | 0.4744 |  |
| FBAR 4-7 | 0.39 | 0.4405 | 0.4505 | 0.4298 | 0.444 | 0.395 | 0.4267 | 0.369 | 0.3568 | 0.4732 |  |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |
| 2 | 0.0391 | 0.0161 | 0.0033 | 0.0066 | 0.0441 | 0.01 | 0.0062 | 0.0595 | 0.0096 | 0.044 |  |
| 3 | 0.1318 | 0.1468 | 0.093 | 0.078 | 0.1694 | 0.1306 | 0.2743 | 0.1794 | 0.1734 | 0.2955 |  |
| 4 | 0.3898 | 0.3238 | 0.4683 | 0.3722 | 0.2313 | 0.3312 | 0.4064 | 0.5606 | 0.3686 | 0.4545 |  |
| 5 | 0.5632 | 0.5044 | 0.3988 | 0.5161 | 0.3632 | 0.3005 | 0.4845 | 0.7115 | 0.5542 | 0.5317 |  |
| 6 | 0.9427 | 0.5955 | 0.4282 | 0.3796 | 0.4122 | 0.3079 | 0.3085 | 0.929 | 0.6375 | 0.6137 |  |
| 7 | 0.6346 | 0.4763 | 0.4332 | 0.424 | 0.3366 | 0.3141 | 0.4012 | 0.7377 | 0.5229 | 0.5367 |  |
| +gp | 0.6346 | 0.4763 | 0.4332 | 0.424 | 0.3366 | 0.3141 | 0.4012 | 0.7377 | 0.5229 | 0.5367 |  |
| FBAR 4-7 | 0.6326 | 0.475 | 0.4321 | 0.423 | 0.3358 | 0.3135 | 0.4002 | 0.7347 | 0.5208 | 0.5341 |  |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| 2 | 0.1139 | 0.1174 | 0.082 | 0.0143 | 0.0248 | 0.0714 | 0.0254 | 0.1098 | 0.0268 | 0.0553 |  |
| 3 | 0.3993 | 0.3569 | 0.2936 | 0.1524 | 0.2985 | 0.2351 | 0.3434 | 0.4126 | 0.3297 | 0.2131 |  |
| 4 | 0.6653 | 0.4861 | 0.4194 | 0.3635 | 0.4433 | 0.5338 | 0.4751 | 0.6119 | 0.4603 | 0.3772 |  |
| 5 | 0.6748 | 0.3824 | 0.6113 | 0.4318 | 0.4923 | 0.5535 | 0.5109 | 0.5136 | 0.7005 | 0.5305 |  |
| 6 | 0.6018 | 0.5149 | 0.3946 | 0.5456 | 0.5459 | 0.5977 | 0.6476 | 0.5204 | 0.3992 | 0.7008 |  |
| 7 | 0.6359 | 0.6188 | 0.7211 | 0.8209 | 0.5534 | 0.4535 | 0.6081 | 0.9265 | 0.4358 | 0.2667 |  |
| +gp | 0.6359 | 0.6188 | 0.7211 | 0.8209 | 0.5534 | 0.4535 | 0.6081 | 0.9265 | 0.4358 | 0.2667 |  |
| FBAR 4-7 | 0.6444 | 0.5006 | 0.5366 | 0.5405 | 0.5087 | 0.5346 | 0.5604 | 0.6431 | 0.4989 | 0.4688 |  |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR 06-08 |
| 2 | 0.0158 | 0.0751 | 0.0339 | 0.1209 | 0.049 | 0.1891 | 0.0745 | 0.0989 | 0.0499 | 0.0615 | 0.0701 |
| 3 | 0.2374 | 0.2899 | 0.1539 | 0.6335 | 0.5247 | 0.519 | 0.3282 | 0.4208 | 0.2889 | 0.3429 | 0.3508 |
| 4 | 0.342 | 0.3088 | 0.498 | 0.5681 | 0.3825 | 0.6877 | 0.4956 | 0.3173 | 0.3844 | 0.3814 | 0.361 |
| 5 | 0.2533 | 0.2623 | 0.6803 | 0.4746 | 0.3138 | 0.9087 | 0.5324 | 0.3706 | 0.237 | 0.3039 | 0.3038 |
| 6 | 0.6827 | 0.3146 | 0.36 | 0.3505 | 0.3263 | 0.3788 | 0.7043 | 0.3427 | 0.2633 | 0.2016 | 0.2692 |
| 7 | 1.033 | 0.4899 | 0.1843 | 0.1447 | 0.1474 | 0.2796 | 0.3302 | 0.4031 | 0.4117 | 0.2347 | 0.3498 |
| +gp | 1.033 | 0.4899 | 0.1843 | 0.1447 | 0.1474 | 0.2796 | 0.3302 | 0.4031 | 0.4117 | 0.2347 |  |
| FBAR 4-7 | 0.5778 | 0.3439 | 0.4307 | 0.3845 | 0.2925 | 0.5637 | 0.5156 | 0.3584 | 0.3241 | 0.2804 |  |

Table 6.8.10. Sole in VIIa. Stock numbers-at-age (start of year, in thousands).

|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 3695 | 10180 | 3187 | 13142 | 5876 | 6690 | 3862 | 15821 | 9090 | 8947 |  |  |  |
| 3 | 8350 | 3316 | 9104 | 2854 | 11542 | 5293 | 5804 | 3467 | 14105 | 8163 |  |  |  |
| 4 | 4145 | 6704 | 2588 | 7597 | 2237 | 9596 | 4092 | 4895 | 2741 | 11851 |  |  |  |
| 5 | 1368 | 2791 | 4071 | 1647 | 4786 | 1477 | 6413 | 2435 | 3200 | 1863 |  |  |  |
| 6 | 4389 | 794 | 1451 | 2222 | 961 | 2701 | 823 | 3586 | 1467 | 1936 |  |  |  |
| 7 | 939 | 2586 | 498 | 802 | 1235 | 505 | 1643 | 510 | 2231 | 907 |  |  |  |
| +gp | 8213 | 5535 | 4321 | 3419 | 2830 | 3223 | 2224 | 2196 | 2046 | 1718 |  |  |  |
| TOTAL | 31100 | 31906 | 25219 | 31683 | 29466 | 29484 | 24861 | 32910 | 34881 | 35386 |  |  |  |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |
| 2 | 5132 | 4599 | 2563 | 5898 | 15874 | 16320 | 23828 | 3439 | 3535 | 4370 |  |  |  |
| 3 | 7993 | 4466 | 4095 | 2312 | 5302 | 13744 | 14620 | 21426 | 2932 | 3168 |  |  |  |
| 4 | 6409 | 6339 | 3489 | 3376 | 1935 | 4049 | 10913 | 10056 | 16203 | 2231 |  |  |  |
| 5 | 7458 | 3928 | 4149 | 1977 | 2106 | 1389 | 2631 | 6577 | 5194 | 10140 |  |  |  |
| 6 | 897 | 3842 | 2146 | 2520 | 1067 | 1325 | 931 | 1466 | 2921 | 2700 |  |  |  |
| 7 | 1146 | 316 | 1917 | 1265 | 1560 | 640 | 881 | 619 | 524 | 1397 |  |  |  |
| +gp | 2548 | 2390 | 1182 | 2148 | 2042 | 2873 | 2687 | 1617 | 1326 | 642 |  |  |  |
| TOTAL | 31583 | 25880 | 19541 | 19495 | 29885 | 40340 | 56492 | 45200 | 32635 | 24649 |  |  |  |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |
| 2 | 5509 | 12495 | 4846 | 6129 | 5231 | 2014 | 2514 | 7978 | 6648 | 5887 |  |  |  |
| 3 | 3784 | 4449 | 10054 | 4039 | 5467 | 4617 | 1696 | 2217 | 6468 | 5857 |  |  |  |
| 4 | 2133 | 2296 | 2817 | 6782 | 3138 | 3670 | 3303 | 1089 | 1328 | 4209 |  |  |  |
| 5 | 1281 | 992 | 1278 | 1676 | 4267 | 1823 | 1947 | 1858 | 534 | 758 |  |  |  |
| 6 | 5392 | 590 | 613 | 627 | 985 | 2360 | 948 | 1057 | 1006 | 240 |  |  |  |
| 7 | 1323 | 2673 | 319 | 374 | 329 | 516 | 1175 | 449 | 568 | 611 |  |  |  |
| +gp | 1035 | 1040 | 1889 | 1279 | 1290 | 1005 | 796 | 917 | 1254 | 1039 |  |  |  |
| TOTAL | 20457 | 24536 | 21815 | 20906 | 20706 | 16005 | 12379 | 15565 | 17807 | 18600 |  |  |  |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | GMST 70-07 | AMST 70-07 |
| 2 | 5905 | 3882 | 2771 | 3038 | 3212 | 3161 | 1684 | 1999 | 2158 | 2397 | 0 | 5326 | 6604 |
| 3 | 5040 | 5260 | 3258 | 2424 | 2436 | 2768 | 2367 | 1414 | 1638 | 1858 | 2044 | 4812 | 5939 |
| 4 | 4282 | 3597 | 3562 | 2528 | 1164 | 1304 | 1490 | 1543 | 840 | 1111 | 1197 | 3586 | 4518 |
| 5 | 2612 | 2752 | 2390 | 1959 | 1296 | 719 | 593 | 822 | 1016 | 518 | 690 | 2160 | 2767 |
| 6 | 404 | 1834 | 1916 | 1095 | 1102 | 857 | 262 | 315 | 513 | 726 | 348 | 1256 | 1625 |
| 7 | 108 | 185 | 1212 | 1209 | 698 | 720 | 531 | 117 | 202 | 357 | 542 | 722 | 927 |
| +gp | 362 | 643 | 742 | 1566 | 597 | 1066 | 932 | 669 | 354 | 291 | 467 |  |  |
| TOTAL | 18714 | 18153 | 15851 | 13819 | 10506 | 10594 | 7859 | 6879 | 6723 | 7256 | 5288 |  |  |

Table 6.8.11. Sole in VIIa. Summary.

|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 2 |  |  |  |  |  |
| 1970 | 3695 | 6709 | 6071 | 1785 | 0.294 | 0.39 |
| 1971 | 10180 | 6982 | 5895 | 1882 | 0.3192 | 0.4405 |
| 1972 | 3187 | 5277 | 4653 | 1450 | 0.3117 | 0.4505 |
| 1973 | 13142 | 6142 | 4831 | 1428 | 0.2956 | 0.4298 |
| 1974 | 5876 | 5697 | 4707 | 1307 | 0.2777 | 0.444 |
| 1975 | 6690 | 5704 | 4963 | 1441 | 0.2903 | 0.395 |
| 1976 | 3862 | 5036 | 4509 | 1463 | 0.3245 | 0.4267 |
| 1977 | 15821 | 4610 | 3947 | 1147 | 0.2906 | 0.369 |
| 1978 | 9090 | 5383 | 4498 | 1106 | 0.2459 | 0.3568 |
| 1979 | 8947 | 6327 | 5241 | 1614 | 0.3079 | 0.4732 |
| 1980 | 5132 | 6078 | 5219 | 1941 | 0.3719 | 0.6326 |
| 1981 | 4599 | 5641 | 4931 | 1667 | 0.3381 | 0.475 |
| 1982 | 2563 | 4342 | 4003 | 1338 | 0.3342 | 0.4321 |
| 1983 | 5898 | 4818 | 3996 | 1169 | 0.2925 | 0.423 |
| 1984 | 15874 | 6555 | 4497 | 1058 | 0.2353 | 0.3358 |
| 1985 | 16320 | 7257 | 5335 | 1146 | 0.2148 | 0.3135 |
| 1986 | 23828 | 8634 | 6464 | 1995 | 0.3086 | 0.4002 |
| 1987 | 3439 | 8069 | 6803 | 2808 | 0.4128 | 0.7347 |
| 1988 | 3535 | 5781 | 5366 | 1999 | 0.3725 | 0.5208 |
| 1989 | 4370 | 4871 | 4397 | 1833 | 0.4169 | 0.5341 |
| 1990 | 5509 | 3982 | 3383 | 1583 | 0.468 | 0.6444 |
| 1991 | 12495 | 4015 | 2925 | 1212 | 0.4144 | 0.5006 |
| 1992 | 4846 | 4036 | 3136 | 1259 | 0.4014 | 0.5366 |
| 1993 | 6129 | 3324 | 2850 | 1023 | 0.3589 | 0.5405 |
| 1994 | 5231 | 4581 | 3710 | 1374 | 0.3704 | 0.5087 |
| 1995 | 2014 | 3468 | 3101 | 1266 | 0.4083 | 0.5346 |
| 1996 | 2514 | 2760 | 2432 | 1002 | 0.412 | 0.5604 |
| 1997 | 7978 | 2975 | 2193 | 1003 | 0.4574 | 0.6431 |
| 1998 | 6648 | 3865 | 2753 | 911 | 0.331 | 0.4989 |
| 1999 | 5887 | 3973 | 3021 | 863 | 0.2857 | 0.4688 |
| 2000 | 5905 | 3437 | 2716 | 818 | 0.3012 | 0.5778 |
| 2001 | 3882 | 4325 | 3585 | 1053 | 0.2938 | 0.3439 |
| 2002 | 2771 | 3833 | 3382 | 1090 | 0.3223 | 0.4307 |
| 2003 | 3038 | 3372 | 3014 | 1014 | 0.3363 | 0.3845 |
| 2004 | 3212 | 2855 | 2377 | 709 | 0.2983 | 0.2925 |
| 2005 | 3161 | 2588 | 2029 | 855 | 0.4214 | 0.5637 |
| 2006 | 1684 | 1988 | 1618 | 569 | 0.3517 | 0.5156 |
| 2007 | 1999 | 1856 | 1492 | 492 | 0.3298 | 0.3584 |
| 2008 | 2158 | 1569 | 1214 | 332 | 0.2735 | 0.3241 |
| 2009 | 2397 | 1513 | 1183 | 324 | 0.2739 | 0.2804 |
| Arith. |  |  |  |  |  |  |
| Mean | 6388 | 4606 | 3811 | 1258 | 0.3341 | 0.4621 |
| 0 Units | (Thousand | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 6.8.12. Sole in VIIa. Input to RCT3 (XSA = XSA estimates at age 2, M2 = abundance indices at age 2 from UK (E\&W) March beam trawl survey, $S 2=$ abundance indices at age 2 from UK(E\&W) September beam trawl survey, M1 and S1 similar as previous but at age 1).

Irish Sea sole recruits - age 2

| 4 | 41 | 2 |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1968 | 3695 | -11 | -11 | -11 | -11 |
| 1969 | 10180 | -11 | -11 | -11 | -11 |
| 1970 | 3187 | -11 | -11 | -11 | -11 |
| 1971 | 13142 | -11 | -11 | -11 | -11 |
| 1972 | 5876 | -11 | -11 | -11 | -11 |
| 1973 | 6690 | -11 | -11 | -11 | -11 |
| 1974 | 3862 | -11 | -11 | -11 | -11 |
| 1975 | 15821 | -11 | -11 | -11 | -11 |
| 1976 | 9090 | -11 | -11 | -11 | -11 |
| 1977 | 8947 | -11 | -11 | -11 | -11 |
| 1978 | 5132 | -11 | -11 | -11 | -11 |
| 1979 | 4599 | -11 | -11 | -11 | -11 |
| 1980 | 2563 | -11 | -11 | -11 | -11 |
| 1981 | 5898 | -11 | -11 | -11 | -11 |
| 1982 | 15874 | -11 | -11 | -11 | -11 |
| 1983 | 16320 | -11 | -11 | -11 | -11 |
| 1984 | 23828 | -11 | -11 | -11 | -11 |
| 1985 | 3439 | -11 | -11 | -11 | -11 |
| 1986 | 3535 | -11 | 196 | -11 | -11 |
| 1987 | 4370 | -11 | 304 | -11 | 118 |
| 1988 | 5509 | -11 | 534 | -11 | 218 |
| 1989 | 12495 | -11 | 1286 | -11 | 1712 |
| 1990 | 4846 | -11 | 309 | -11 | 148 |
| 1991 | 6129 | 265 | 330 | -11 | 220 |
| 1992 | 5231 | 307 | 408 | 14 | 83 |
| 1993 | 2014 | 76 | 154 | 7 | 60 |
| 1994 | 2514 | 85 | 126 | 19 | 246 |
| 1995 | 7978 | 343 | 577 | 485 | 886 |
| 1996 | 6648 | 324 | 716 | 107 | 1158 |
| 1997 | 5887 | 174 | 293 | 36 | 539 |
| 1998 | 5905 | -11 | 464 | 34 | 385 |
| 1999 | 3882 | -11 | 284 | -11 | 354 |
| 2000 | 2771 | -11 | 61 | -11 | 91 |
| 2001 | 3038 | -11 | 210 | -11 | 205 |
| 2002 | 3212 | -11 | 240 | -11 | 242 |
| 2003 | 3161 | -11 | 165 | -11 | 406 |
| 2004 | 1684 | -11 | 110 | -11 | 53 |
| 2005 | 1999 | -11 | 93 | -11 | 107 |
| 2006 | -11 | -11 | 125 | -11 | 125 |
| 2007 | -11 | -11 | 150 | -11 | 126 |
| 2008 | -11 | -11 | -11 | -11 | 57 |
|  |  |  |  |  |  |

M2
S2
M1

## Table 6.8.13. Sole in VIIa.

```
Analysis by RCT3 ver3.1 of data from file
SOL7aRCT.txt
lrish Sea sole recruits - age 2
Data for 4 surveys over 40 years: 1969-2008
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates shrunk towards mean
Minimum S.E. for any survey taken as .00
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
Yearclass = 2007
    |---------Regression-----------------------------
Survey/ Slope Inter- Std Rsquare No. Index Predicted Std WAP
Series cept Error Pts Value Value Error Weights
M2 
M1 
                                    VPA Mean = 8.59 . 650 . .114
Yearclass = 2008
    1---------Regression------------------------------
Survey/Slope Inter- Std Rsquare No. Index Predicted Std WAP
Series cept Error Pts Value Value Error Weights
M2
M1
S1 .71 4.44 .46 . 581 19 4.06 7.32 .524 . 606 Predicted value = 7.32= 1510 60.6% weight
                VPA Mean = 8.59 . 650 . 394
Predicted value \(=7.32=\quad 1510\)
\(39.4 \%\) weight
Predicted value \(=8.59=\quad 5378\)
39.4 \% weight
Year Weighted Log Int Ext Var VPA Log Class Average WAP Std Std Ratio VPA
Prediction Error Error
\(2007 \quad 2877 \quad 7.96 \quad .22 \quad .16 \quad .52\)
\(2008 \quad 2489 \quad 7.82 \quad .41 \quad .62 \quad 2.32\)
```

Table 6.8.14. VIIa sole : Catch forecast input data.
MFDP version 1a
Run: s71
Time and date: 16:01 31/05/2010
Fbar age range: 4-7


| $\begin{array}{\|l} 2011 \\ \text { Age } \\ \hline \end{array}$ | $N$ | M |  |  | PF |  | SWt |  | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 |  | 5326 | 0.1 | 0.38 |  | 0 | 0 | 0.177333 | 0.0701 | 0.190067 |
| 3 | . |  | 0.1 | 0.71 |  | 0 | 0 | 0.209667 | 0.3508 | 0.233367 |
| 4 | . |  | 0.1 | 0.97 |  | 0 | 0 | 0.253 | 0.361 | 0.263333 |
| 5 | . |  | 0.1 | 0.98 |  | 0 | 0 | 0.29 | 0.3038 | 0.301300 |
| 6 | . |  | 0.1 | 1 |  | 0 | 0 | 0.304333 | 0.2692 | 0.308267 |
| 7 |  |  | 0.1 | 1 |  | 0 | 0 | 0.324333 | 0.3498 | 0.327567 |
| 8 | . |  | 0.1 | 1 |  | 0 | 0 | 0.381616 | 0.3498 | 0.391602 |



Input units are thousands and kg - output in tonnes

Table 6.8.15.a. VIIa sole: management option table; status quo forecast unscaled.

| MFDP version 1a <br> Run: s7a <br> IRISH SEA SOLE <br> Time and date: 15:50 31/05/2010 <br> Fbar age range: 4-7 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 <br> Biomass | SSB | FMult | FBar | Landings |  |  |
| 1902 | 1448 | 1.0000 | 0.3210 | 439 |  |  |
| $2011$ <br> Biomass | SSB | FMult | FBar | Landings | $2012$ <br> Biomass | SSB |
| 1962 | 1527 | 0.0000 | 0.0000 | 0 | 2962 | 2216 |
| . | 1527 | 0.1000 | 0.0321 | 54 | 2907 | 2163 |
| . | 1527 | 0.2000 | 0.0642 | 107 | 2854 | 2112 |
| . | 1527 | 0.3000 | 0.0963 | 158 | 2803 | 2062 |
| . | 1527 | 0.4000 | 0.1284 | 207 | 2753 | 2014 |
| . | 1527 | 0.5000 | 0.1605 | 254 | 2705 | 1967 |
| . | 1527 | 0.6000 | 0.1926 | 301 | 2658 | 1922 |
| . | 1527 | 0.7000 | 0.2247 | 345 | 2613 | 1879 |
| . | 1527 | 0.8000 | 0.2568 | 389 | 2569 | 1837 |
| . | 1527 | 0.9000 | 0.2889 | 431 | 2527 | 1796 |
| . | 1527 | 1.0000 | 0.3210 | 472 | 2486 | 1756 |
| . | 1527 | 1.1000 | 0.3530 | 511 | 2446 | 1718 |
| . | 1527 | 1.2000 | 0.3851 | 549 | 2408 | 1681 |
| . | 1527 | 1.3000 | 0.4172 | 586 | 2370 | 1645 |
| . | 1527 | 1.4000 | 0.4493 | 622 | 2334 | 1610 |
| . | 1527 | 1.5000 | 0.4814 | 657 | 2299 | 1577 |
| . | 1527 | 1.6000 | 0.5135 | 691 | 2265 | 1544 |
| . | 1527 | 1.7000 | 0.5456 | 724 | 2233 | 1513 |
| . | 1527 | 1.8000 | 0.5777 | 755 | 2201 | 1482 |
| . | 1527 | 1.9000 | 0.6098 | 786 | 2170 | 1453 |
| . | 1527 | 2.0000 | 0.6419 | 816 | 2140 | 1424 |

Table 6.8.15.b. VIIa sole: management option table; scaled (advice basis).


Table 6.8.16. VIIa sole : forecast detailed results; status quo projection.

| MFDP version 1a <br> Run: s71 <br> Time and date: 16:01 31/05/2010 <br> Fbar age range: 4-7 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year: Age | $\begin{aligned} & 2010 \\ & \mathrm{~F} \\ & \hline \end{aligned}$ |  | F multiplier: CatchNos | 1 <br> Yield |  | Fbar: <br> StockNos | 0.321 <br> Biomass |  | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 |  | 0.0701 | 160 |  | 30 | 2489 |  | 441 | 946 | 168 | 946 | 168 |
| 3 |  | 0.3508 | 577 |  | 135 | 2044 |  | 429 | 1451 | 304 | 1451 | 304 |
| 4 |  | 0.361 | 346 |  | 91 | 1197 |  | 303 | 1161 | 294 | 1161 | 294 |
| 5 |  | 0.3038 | 172 |  | 52 | 690 |  | 200 | 676 | 196 | 676 | 196 |
| 6 |  | 0.2692 | 78 |  | 24 | 348 |  | 106 | 348 | 106 | 348 | 106 |
| 7 |  | 0.3498 | 153 |  | 50 | 542 |  | 176 | 542 | 176 | 542 | 176 |
| 8 |  | 0.3498 | 132 |  | 52 | 467 |  | 178 | 467 | 178 | 467 | 178 |
| Total |  |  | 1619 |  | 434 | 7777 |  | 1833 | 5591 | 1422 | 5591 | 1422 |
| Year: <br> Age | $\begin{aligned} & 2011 \\ & \mathrm{~F} \end{aligned}$ |  | F multiplier: CatchNos | $\begin{aligned} & 1 \\ & \text { Yield } \\ & \hline \end{aligned}$ |  | Fbar: <br> StockNos | 0.321 <br> Biomass |  | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 |  | 0.0701 | 343 |  | 65 | 5326 |  | 944 | 2024 | 359 | 2024 | 359 |
| 3 |  | 0.3508 | 593 |  | 138 | 2100 |  | 440 | 1491 | 313 | 1491 | 313 |
| 4 |  | 0.361 | 377 |  | 99 | 1302 |  | 329 | 1263 | 320 | 1263 | 320 |
| 5 |  | 0.3038 | 189 |  | 57 | 755 |  | 219 | 740 | 215 | 740 | 215 |
| 6 |  | 0.2692 | 104 |  | 32 | 461 |  | 140 | 461 | 140 | 461 | 140 |
| 7 |  | 0.3498 | 68 |  | 22 | 241 |  | 78 | 241 | 78 | 241 | 78 |
| 8 |  | 0.3498 | 181 |  | 71 | 643 |  | 246 | 643 | 246 | 643 | 246 |
| Total |  |  | 1854 |  | 485 | 10828 |  | 2397 | 6862 | 1669 | 6862 | 1669 |
| Year: <br> Age | $\begin{aligned} & 2012 \\ & \mathrm{~F} \\ & \hline \end{aligned}$ |  | F multiplier: CatchNos | 1 <br> Yield |  | Fbar: StockNos | 0.321 <br> Biomass |  | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 2 |  | 0.0701 | 343 |  | 65 | 5326 |  | 944 | 2024 | 359 | 2024 | 359 |
| 3 |  | 0.3508 | 1269 |  | 296 | 4493 |  | 942 | 3190 | 669 | 3190 | 669 |
| 4 |  | 0.361 | 387 |  | 102 | 1338 |  | 338 | 1298 | 328 | 1298 | 328 |
| 5 |  | 0.3038 | 205 |  | 62 | 821 |  | 238 | 805 | 233 | 805 | 233 |
| 6 |  | 0.2692 | 113 |  | 35 | 504 |  | 153 | 504 | 153 | 504 | 153 |
| 7 |  | 0.3498 | 90 |  | 29 | 319 |  | 103 | 319 | 103 | 319 | 103 |
| 8 |  | 0.3498 | 159 |  | 62 | 564 |  | 215 | 564 | 215 | 564 | 215 |
| Total |  |  | 2566 |  | 652 | 13364 |  | 2935 | 8703 | 2061 | 8703 | 2061 |

Table 6.8.17. Sole VIIa. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.


Table 6.8.18. Sole VIIa: Yield-per-recruit.
MFYPR version 2 a
Run: s7a
Time and date: 16:09 31/05/2010
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 10.5083 | 3.3906 | 9.5866 | 3.2151 | 9.5866 | 3.2151 |
| 0.1000 | 0.0321 | 0.2346 | 0.0783 | 8.1649 | 2.5165 | 7.2470 | 2.3420 | 7.2470 | 2.3420 |
| 0.2000 | 0.0642 | 0.3736 | 0.1208 | 6.7773 | 2.0056 | 5.8633 | 1.8320 | 5.8633 | 1.8320 |
| 0.3000 | 0.0963 | 0.4659 | 0.1462 | 5.8574 | 1.6716 | 4.9471 | 1.4989 | 4.9471 | 1.4989 |
| 0.4000 | 0.1284 | 0.5317 | 0.1625 | 5.2015 | 1.4370 | 4.2947 | 1.2652 | 4.2947 | 1.2652 |
| 0.5000 | 0.1605 | 0.5812 | 0.1734 | 4.7094 | 1.2637 | 3.8061 | 1.0927 | 3.8061 | 1.0927 |
| 0.6000 | 0.1926 | 0.6197 | 0.1809 | 4.3261 | 1.1308 | 3.4261 | 0.9605 | 3.4261 | 0.9605 |
| 0.7000 | 0.2247 | 0.6507 | 0.1862 | 4.0188 | 1.0258 | 3.1221 | 0.8564 | 3.1221 | 0.8564 |
| 0.8000 | 0.2568 | 0.6761 | 0.1900 | 3.7669 | 0.9411 | 2.8734 | 0.7724 | 2.8734 | 0.7724 |
| 0.9000 | 0.2889 | 0.6974 | 0.1928 | 3.5566 | 0.8715 | 2.6661 | 0.7035 | 2.6661 | 0.7035 |
| 1.0000 | 0.3210 | 0.7155 | 0.1948 | 3.3783 | 0.8132 | 2.4907 | 0.6459 | 2.4907 | 0.6459 |
| 1.1000 | 0.3530 | 0.7310 | 0.1964 | 3.2252 | 0.7639 | 2.3405 | 0.5973 | 2.3405 | 0.5973 |
| 1.2000 | 0.3851 | 0.7445 | 0.1975 | 3.0923 | 0.7218 | 2.2105 | 0.5558 | 2.2105 | 0.5558 |
| 1.3000 | 0.4172 | 0.7564 | 0.1983 | 2.9760 | 0.6853 | 2.0969 | 0.5200 | 2.0969 | 0.5200 |
| 1.4000 | 0.4493 | 0.7669 | 0.1989 | 2.8732 | 0.6535 | 1.9968 | 0.4888 | 1.9968 | 0.4888 |
| 1.5000 | 0.4814 | 0.7763 | 0.1994 | 2.7818 | 0.6255 | 1.9080 | 0.4614 | 1.9080 | 0.4614 |
| 1.6000 | 0.5135 | 0.7847 | 0.1997 | 2.7001 | 0.6008 | 1.8287 | 0.4373 | 1.8287 | 0.4373 |
| 1.7000 | 0.5456 | 0.7923 | 0.2000 | 2.6265 | 0.5788 | 1.7576 | 0.4158 | 1.7576 | 0.4158 |
| 1.8000 | 0.5777 | 0.7991 | 0.2001 | 2.5599 | 0.5591 | 1.6934 | 0.3967 | 1.6934 | 0.3967 |
| 1.9000 | 0.6098 | 0.8054 | 0.2002 | 2.4994 | 0.5414 | 1.6353 | 0.3795 | 1.6353 | 0.3795 |
| 2.0000 | 0.6419 | 0.8111 | 0.2003 | 2.4441 | 0.5254 | 1.5824 | 0.3641 | 1.5824 | 0.3641 |

Reference poin:- multiplie Absolute F

| Fbar(4-7) | 1 | 0.321 |
| :--- | ---: | ---: |
| FMax | 2.13640 | 0.6857 |
| F0.1 | 0.4471 | 0.1435 |
| F35\%SPR | 0.4788 | 0.1537 |

Weights in kilograms

Table 6.8.19. Sole VIIa: SUM file input to MSY explorations in SRMSYMS.

| Stock summary, Sole ,Irish Sea 12 |  |  | ,17: 1, 31/5/2010 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 100 |  |  |  |  |  |  |  |  |  |  |  |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| Recruits, age 2, (thousands) |  |  |  |  |  |  |  |  |  |  |  |
| 21000 |  |  |  |  |  |  |  |  |  |  |  |
| SSB, (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| TSB, (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Catch, Total (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Catch, H.cons (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Mean F, Total |  |  |  |  |  |  |  |  |  |  |  |
| Mean F, H.cons. |  |  |  |  |  |  |  |  |  |  |  |
| 47 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 00 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 00 |  |  |  |  |  |  |  |  |  |  |  |
| 1970 | 3695 | 6070.9 | 6708.5 | 1785 | 1785 | 0 | 0 | 0.39 | 0.39 | 0 | 0 |
| 1971 | 10180 | 5895.4 | 6981.9 | 1882 | 1882 | 0 | 0 | 0.44 | 0.44 | 0 | 0 |
| 1972 | 3187 | 4652.7 | 5277 | 1450 | 1450 | 0 | 0 | 0.45 | 0.45 | 0 | 0 |
| 1973 | 13142 | 4830.9 | 6141.7 | 1428 | 1428 | 0 | 0 | 0.43 | 0.43 | 0 | 0 |
| 1974 | 5876 | 4707.1 | 5697.2 | 1307 | 1307 | 0 | 0 | 0.444 | 0.444 | 0 | 0 |
| 1975 | 6690 | 4963.2 | 5704.7 | 1441 | 1441 | 0 | 0 | 0.395 | 0.395 | 0 | 0 |
| 1976 | 3862 | 4508.5 | 5035.7 | 1463 | 1463 | 0 | 0 | 0.427 | 0.427 | 0 | 0 |
| 1977 | 15821 | 3946.7 | 4610.1 | 1147 | 1147 | 0 | 0 | 0.369 | 0.369 | 0 | 0 |
| 1978 | 9090 | 4497.8 | 5383.2 | 1106 | 1106 | 0 | 0 | 0.357 | 0.357 | 0 | 0 |
| 1979 | 8947 | 5241.2 | 6326.8 | 1614 | 1614 | 0 | 0 | 0.473 | 0.473 | 0 | 0 |
| 1980 | 5132 | 5218.7 | 6077.7 | 1941 | 1941 | 0 | 0 | 0.633 | 0.633 | 0 | 0 |
| 1981 | 4599 | 4930.7 | 5641.3 | 1667 | 1667 | 0 | 0 | 0.475 | 0.475 | 0 | 0 |
| 1982 | 2563 | 4003.1 | 4342 | 1338 | 1338 | 0 | 0 | 0.432 | 0.432 | 0 | 0 |
| 1983 | 5898 | 3996.4 | 4817.7 | 1169 | 1169 | 0 | 0 | 0.423 | 0.423 | 0 | 0 |
| 1984 | 15874 | 4497.3 | 6555.1 | 1058 | 1058 | 0 | 0 | 0.336 | 0.336 | 0 | 0 |
| 1985 | 16320 | 5334.8 | 7257.3 | 1146 | 1146 | 0 | 0 | 0.313 | 0.313 | 0 | 0 |
| 1986 | 23828 | 6464.3 | 8634.3 | 1995 | 1995 | 0 | 0 | 0.4 | 0.4 | 0 | 0 |
| 1987 | 3439 | 6802.6 | 8068.8 | 2808 | 2808 | 0 | 0 | 0.735 | 0.735 | 0 | 0 |
| 1988 | 3535 | 5366.5 | 5781 | 1999 | 1999 | 0 | 0 | 0.521 | 0.521 | 0 | 0 |
| 1989 | 4370 | 4396.9 | 4870.3 | 1833 | 1833 | 0 | 0 | 0.534 | 0.534 | 0 | 0 |
| 1990 | 5509 | 3382.7 | 3981.9 | 1583 | 1583 | 0 | 0 | 0.644 | 0.644 | 0 | 0 |
| 1991 | 12495 | 2924.8 | 4015.1 | 1212 | 1212 | 0 | 0 | 0.501 | 0.501 | 0 | 0 |
| 1992 | 4846 | 3136.6 | 4036.4 | 1259 | 1259 | 0 | 0 | 0.537 | 0.537 | 0 | 0 |
| 1993 | 6129 | 2850.1 | 3324.2 | 1023 | 1023 | 0 | 0 | 0.54 | 0.54 | 0 | 0 |
| 1994 | 5231 | 3710.2 | 4580.9 | 1374 | 1374 | 0 | 0 | 0.509 | 0.509 | 0 | 0 |
| 1995 | 2014 | 3100.9 | 3468 | 1266 | 1266 | 0 | 0 | 0.535 | 0.535 | 0 | 0 |
| 1996 | 2514 | 2431.7 | 2759.8 | 1002 | 1002 | 0 | 0 | 0.56 | 0.56 | 0 | 0 |
| 1997 | 7978 | 2192.6 | 2975.4 | 1003 | 1003 | 0 | 0 | 0.643 | 0.643 | 0 | 0 |
| 1998 | 6648 | 2752.4 | 3864.8 | 911 | 911 | 0 | 0 | 0.499 | 0.499 | 0 | 0 |
| 1999 | 5887 | 3020.6 | 3972.7 | 863 | 863 | 0 | 0 | 0.469 | 0.469 | 0 | 0 |
| 2000 | 5905 | 2716.1 | 3436.6 | 818 | 818 | 0 | 0 | 0.578 | 0.578 | 0 | 0 |
| 2001 | 3882 | 3584.6 | 4325.5 | 1053 | 1053 | 0 | 0 | 0.344 | 0.344 | 0 | 0 |
| 2002 | 2771 | 3381.8 | 3833.2 | 1090 | 1090 | 0 | 0 | 0.431 | 0.431 | 0 | 0 |
| 2003 | 3038 | 3014.5 | 3371.9 | 1013.9 | 1013.9 | 0 | 0 | 0.384 | 0.384 | 0 | 0 |
| 2004 | 3212 | 2376.4 | 2854.4 | 709 | 709 | 0 | 0 | 0.293 | 0.293 | 0 | 0 |
| 2005 | 3161 | 2029.1 | 2588 | 855 | 855 | 0 | 0 | 0.564 | 0.564 | 0 | 0 |
| 2006 | 1684 | 1617.8 | 1988.4 | 569 | 569 | 0 | 0 | 0.516 | 0.516 | 0 | 0 |
| 2007 | 1999 | 1491.6 | 1855.5 | 492 | 492 | 0 | 0 | 0.358 | 0.358 | 0 | 0 |
| 2008 | 2158 | 1213.8 | 1568.8 | 332 | 332 | 0 | 0 | 0.324 | 0.324 | 0 | 0 |
| 2009 | 2397 | 1183.3 | 1513.8 | 324 | 324 | 0 | 0 | 0.28 | 0.28 | 0 | 0 |

Table 6.8.20. Sole VIIa: SEN file input to MSY explorations in SRMSYMS.

| Input to sensitivity analysis, SOL,7A |  |
| :---: | :---: |
| 2, 8, 2010, 3 |  |
| 1, 0, 0 |  |
| 'N2', | 5326, 0.64 |
| 'N3' , | 2044, 0.35 |
| 'N4' , | 1196, 0.23 |
| 'N5' , | 690, 0.20 |
| 'N6' | 348, 0.20 |
| 'N7' | 542, 0.17 |
| 'N8' , | 466, 0.17 |
| 'sH2', | 0.070, 0.28 |
| 'sH3', | 0.351, 0.16 |
| 'sH4', | 0.361, 0.21 |
| 'sH5', | 0.304, 0.20 |
| 'sH6', | 0.269, 0.14 |
| 'sH7', | 0.350, 0.20 |
| 'sH8' , | 0.350, 0.20 |
| 'WH2', | 0.190, 0.17 |
| 'WH3', | 0.233, 0.08 |
| 'WH4', | 0.263, 0.09 |
| 'WH5', | 0.301, 0.06 |
| 'WH6', | 0.308, 0.14 |
| 'WH7', | 0.328, 0.17 |
| 'WH8', | 0.392, 0.16 |
| 'WS2', | 0.177, 0.14 |
| 'WS3', | 0.210, 0.14 |
| 'WS4', | 0.253, 0.07 |
| 'WS5', | 0.290, 0.06 |
| 'WS6', | 0.304, 0.10 |
| 'WS7', | 0.324, 0.13 |
| 'WS8', | 0.382, 0.16 |
| 'M2', | $0.10,0.10$ |
| 'M3' , | 0.10, 0.10 |
| 'M4' , | 0.10, 0.10 |
| 'M5' , | 0.10, 0.10 |
| 'M6' , | 0.10, 0.10 |
| 'M7' , | 0.10, 0.10 |
| 'M8' , | 0.10, 0.10 |
| 'MT2' , | 0.38, 0.10 |
| 'MT3' , | 0.71, 0.10 |
| 'MT4', | 0.97, 0.10 |
| 'MT5', | 0.98, 0.10 |
| 'MT6' , | 1.00, 0.10 |
| 'MT7', | 1.00, 0.00 |
| 'MT8' , | 1.00, 0.00 |
| 'R11', | 5326, 0.64 |
| 'R12', | 5326, 0.64 |
| 'HF10', | 1, 0.12 |
| 'HF11', | 1, 0.12 |
| 'HF12', | 1, 0.12 |
| 'K10', | 1, 0.10 |
| 'K11', | 1, 0.10 |
| 'K12', | 1, 0.10 |
| Sole |  |
| Irish Sea |  |
|  | 1 |
| 「281 |  |
|  | 1 |
| H.cons. |  |
| 47 |  |
| 1970 | 2009 |
| Stock numbers in 2010 are VPA survivors$-1$ |  |

Table 6.8.21. Sole VIIa: Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships and the yield-per-recruit $\mathrm{F}_{\text {msy }}$ proxies.

Stock name
Sol-7a
Sen filename
sol7a.sen
pf, pm
0
Number of iterations

$$
1000
$$

Simulate variation in Biological parameters TRUE
SR relationship constrained TRUE

## Ricker

691/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 0.48 | 0.18 | 6,483 | 1,174 | 0.63 | 0.72 | 2.18 | 0.000118357 | 75.04 |
| Mean | 0.47 | 0.17 | 37,151 | 4,372 | 0.65 | 0.71 | 2.29 | 0.000117529 |  |
| 5\%ile | 0.22 | 0.09 | 3,318 | 702 | 0.48 | 0.16 | 1.49 | 2.67E-05 |  |
| 25\%ile | 0.30 | 0.12 | 4,732 | 1,000 | 0.57 | 0.42 | 1.78 | 6.87E-05 |  |
| 50\%ile | 0.40 | 0.16 | 6,866 | 1,292 | 0.65 | 0.66 | 2.13 | 0.000108706 |  |
| 75\%ile | 0.53 | 0.20 | 10,611 | 1,667 | 0.72 | 0.95 | 2.57 | 0.000156316 |  |
| 95\%ile | 0.94 | 0.30 | 27,814 | 3,535 | 0.82 | 1.48 | 3.70 | 0.000244428 |  |
| CV | 0.63 | 0.38 | 14.07 | 10.85 | 0.17 | 0.55 | 0.31 | 0.55 |  |

## Beverton-Holt

669/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 0.49 | 0.12 | 11,657 | 1,444 | 0.73 | 1.50 | 13865 | 6306 | 75.50 |
| Mean | 0.58 | 0.10 | 49,698 | 2,525 | 0.80 | 1.56 | 27870 | 17232 |  |
| 5\%ile | 0.20 | 0.02 | 3,713 | 661 | 0.12 | 1.23 | 6646 | 1424 |  |
| 25\%ile | 0.31 | 0.07 | 6,802 | 996 | 0.45 | 1.40 | 8961 | 2993 |  |
| 50\%ile | 0.43 | 0.10 | 12,000 | 1,380 | 0.79 | 1.54 | 12894 | 5753 |  |
| 75\%ile | 0.65 | 0.14 | 39,614 | 2,412 | 1.14 | 1.72 | 22390 | 12677 |  |
| 95\%ile | 1.57 | 0.21 | 206,140 | 8,013 | 1.53 | 1.96 | 86895 | 60872 |  |
| CV | 0.87 | 0.55 | 2.40 | 1.46 | 0.55 | 0.15 | 2.07 | 2.57 |  |

Smooth hockeystick
691/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash |  | Fmsy | Bmsy |  | MSY |  | ADMB Alpha ADMB Beta Unscaled Alpha Unscaled Beta AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Deterministic | 0.31 | 0.31 | 4,093 | 1,201 | 0.45 | 1.07 | 0.76 | 4090.96 |
| Mean | 0.28 | 0.24 | 7,253 | 1,256 | 0.47 | 1.11 | 0.78 | 4232.701592 |
| 5\%ile | 0.16 | 0.04 | 3,087 | 824 | 0.38 | 0.77 | 0.63 | 2942.24 |
| 25\%ile | 0.22 | 0.19 | 3,926 | 1,041 | 0.42 | 1.00 | 0.71 | 3792.365 |
| 50\%ile | 0.28 | 0.25 | 4,377 | 1,220 | 0.46 | 1.10 | 0.77 | 4181.98 |
| 75\%ile | 0.32 | 0.30 | 5,142 | 1,429 | 0.50 | 1.21 | 0.84 | 4616.67 |
| 95\%ile | 0.42 | 0.40 | 30,279 | 1,847 | 0.58 | 1.53 | 0.97 | 5829.76 |
| CV | 0.30 | 0.39 | 1.47 | 0.25 | 0.16 | 0.20 | 0.16 | 0.20 |




Figure 6.8.1. Sole in VIIa. Relative cpue and effort-series for beam trawlers from Belgium (B-BT), the UK (UK-BT) and Ireland (IRE-BT); for otter trawlers from the UK (UK-OT) and Ireland (IREOT); and cpue series for the UK (EandW) September beam trawl survey (UK-BTS-Sept).

UK (E_W) SEPTEMBER BEAM TRAWL


UK (E_W) SEPTEMBER BEAM TRAWL


UK-(E+W) MARCH BEAM TRAWL


UK-(E+W) MARCH BEAM TRAWL


Figure 6.8.2. Sole in VIIa. Catchability residuals of final XSA run.


Figure 6.8.3. Sole in VIIa. Summary ( $\mathrm{Blim}_{\mathrm{lim}}=2200 \mathrm{t}, \mathrm{B}_{\mathrm{pa}}=3100 \mathrm{t}, \mathrm{F}_{\mathrm{lim}}=0.4, \mathrm{~F}_{\mathrm{pa}}=0.3$ ).


Figure 6.8.4. Sole in VIIa. Retrospective.


Figure 6.8.5. Sole in VIIa. Comparison of trends in Recruitment, SSB and Fishing Mortality from last year's (WG2009) and this year's final assessment (WG2010, same procedure as last year incl. an additional datayear).


Figure 6.8.6. VIIa Sole : Yield-per-recruit and short-term forecast results.


Table 6.8.7. VIIa Sole : MSY fitted stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $\mathrm{F}_{\text {msy }}$. Right hand panels: curves plotted from the first 100 MCMC re-samples with converged $F_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of Fmsy $^{\text {from the }} \mathbf{1 0 0 0}$ re-samples.

## Sol-7a Beverton-Holt



Table 6.8.8. VIIa Sole : Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Beverton and Holt stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy. }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.


Table 6.8.9. VIIa Sole: Estimates of $F$ reference points and equilibrium yield and SSB against fishing mortality using a Ricker stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

## Sol-7a Smooth hockeystick



Table 6.8.10. VIIa Sole: Estimates of F reference points and equilibrium yield and SSB against fishing mortality using a Hockey Stick and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy. }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

Sol-7a - Per recruit statistics


Table 6.8.11. VIIa Sole: Fitted yield-per-recruit F reference points, yield-per-recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first $\mathbf{1 0 0}$ re-samples.

### 7.1 Celtic Sea overview

There is no overview.

### 7.2 Cod in Division VIIe-k (celtic Sea)

Type of assessment in 2010

## Trends analysis

For Celtic Sea cod, the Benchmark Workshop WKROUND 2009 concluded that more work was required before this stock could be benchmarked. The Review Group of WGCSE 2009 added that shortcomings of the data and reconstruction of datasets should be completed in order to continue using an aged based assessment in future.

The recommendations made by WKROUND 2009 were:

- Improvement of the quality of assessment input data, of documentation on data correction in the Stock Annex and data integration and fishery description at regional level through a regional database.
- Evaluation of sampling levels by fleet required to get precise discard estimates for stock assessment. The RG concurred with the conclusion drawn by the Benchmark that cooperative projects with industry on self sampling, and reference fleets, etc should be developed to obtain better estimates of discards. Datasets obtained through fishers science partnerships should be used to complement those discard data collected by fishery observers.
- Estimates of "true landings" as reported landings data and landings equivalents since 2003 are thought to be underestimated.
- International coordination on maturity sampling as there is evidence that maturity has changed for this stock. A directed survey might be needed.
- Improvement on knowledge on stock structure and migration behaviour.
- Reduction of noise in the data from the surveys.

Solutions to those recommendations have been suggested by WGCSE in last year's Report. No new development has been presented to the Working Group this year. Some effort to improve the knowledge on this stock is currently done through survey and industry-science partnerships. Those initiatives are summarized later in this section.

## ICES advice applicable to 2009

"Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects

The current fishing mortality is estimated at 0.67 , which is well above the range that would lead to high long-term yields and low risk of stock depletion.

Exploitation boundaries in relation to precautionary limits
The exploitation boundaries in relation to precautionary limits imply landings of less than $2600 t$ in 2009, which is expected to rebuild SSB to the Bpa $(=8800 t)$ in 2010.

Conclusion on exploitation boundaries

ICES recommends a $50 \%$ reduction in fishing mortality which is associated with landings in 2009 of $2600 t$; the SSB is then expected to reach $B_{p a}$ in 2010. This fishing mortality also corresponds to high long-term yield and low risk of stock depletion."

## ICES advice applicable to 2010

"ICES advises on the basis of precautionary considerations that fishing effort and catches should be reduced although it is not possible to determine the appropriate scale of such reduction."

### 7.2.1 Genera

## Stock description and management units

The 2010 TAC was set for ICES Areas VIIb-c, VIIe-k, VIII, IX, X, and CECAF 34.1.1(1), excluding VIId. This is more representative of the stock area than in the previous years as the cod population in VIId is more relevant to the North Sea population but landings from VIIbc are not included in the assessment area (see Section 7.3 for these).


Red Boxes-TAC/Management Areas Blue Shading- Assessment Area.
Management applicable to 2009 and 2010

TAC 2009

| Species: Cod Gadus morhua |  | Zone: | VIIb-c, VIIe-k, VIII, IX and X; EC waters of CECAF 34.1.1 (COD/7XAD34) |
| :---: | :---: | :---: | :---: |
| Belgium | 167 |  |  |
| France | 2735 |  |  |
| Ireland | 825 |  |  |
| The Netherlands | 1 |  |  |
| United Kingdom | 295 |  |  |
| EC | 4023 |  |  |
| TAC | 4023 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TAC 2010

| Species:Cod <br> Gadus morhua | Zone: | VIIb, VIIc, VIIe-k, VIII, IX and X; EU waters of CECAF <br> 34.1 .1 <br> (COD/7XAD34) |
| :--- | :--- | :--- | :--- |
| Belgium | 167 |  |
| France | 2735 |  |
| Ireland | 825 |  |
| The Netherlands | 1 | Analytical TAC |
| United Kingdom | 295 |  |
| EU | 4023 |  |
| TAC |  |  |

## Fishery in 2009

Landings data used by the WG are shown in Table 7.2.1. The Irish landings in 2008 were revised upward by 11 t to 1221 t . No revision was required for UK and Belgium.

French landings were very preliminary this year due to some major changes in the administrative processing of the data for all French fisheries. Data compilation showed strong evidence that official available landings were very partial (around $30 \%$ of French quota) and well below the landings reported by the French fishing organizations involved in the cod fishery ( $75 \%$ of the French quota). Quality checks of the industry data for 2009 based on the fishing activity per quarter during the period 2006-2008, suggests the industry data are the best estimates of the French landings for 2009. Subsequent estimates and raising, when possible, were based on the industry dataset. French landings in 2009 will be revised accordingly with official data whenever possible. However, French landings indicate that only $72 \%$ of the national quota has been taken. This is mainly related to 1) decommissioning of a substantial number of vessels in 2009,2 ) low fish market prices for cod landings which led the vessels to direct their fishing effort towards other demersal species (e.g. haddock).

International landings have decreased in 2008 (3600 t) and 2009 ( 3200 t ) after the 2007 peak of 4200 t , which corresponded to approximately half of the average ( 8200 t ) of the time-series. They are now close to their lowest historical values. Since 1988, French landings accounted for $\sim 70 \%$ of the international landings and they have declined to around $58 \%$ of the total in the recent years. Irish landings accounted on average at $14 \%$ but more recently $\sim 28 \%$. UK and Belgium have contributed on average to $9 \%$ and $4 \%$.

There is no information on the absolute level of misreporting for this stock but there is evidence that misreporting has increased from 2002 when quotas became restrictive. Irish landings data in some years have been corrected for area misreporting into the southern rectangles of VIIa. These misreporting estimates are summarized in table below.

| Year | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | ---: | ---: | ---: | ---: | ---: | :---: |
| Mis alloc $(\mathrm{t})$ | 108 | 54 | 103 | 514 | 558 | 55 |

French landings have been corrected with high grading estimates from 2003 to 2005. The method used to estimate the high graded component is described in WD\#1 of the WG SSDS 2006. For smaller length classes, a scaling of French numbers-at-length based on UK length frequencies or UK number-at-length has been used to estimate length compositions of the French component of high grading. The accuracy of this method is unknown but it probably underestimates the high-grading levels for those years. Unfortunately, the sampling level of total catch at sea in that period was too poor to get an estimate of the level of bias.

This method was not applied from 2006 onward because high grading was also observed in the UK landings. Instead, self sampling data obtained in 2008-2009 have been used to estimate the French high grading level, assuming that the discarding practices in 2006-2007 were the same as those observed in 2008 for the main selfsampled fleet. Applying this method back to 2003 was considered inappropriate. The representatives of Fishermen Organisations at WKROUND 2009 indicated that the discarding level was probably not the same in earlier years as high-grading practices are linked to the level of the TAC. The whole method has been described in the WD\#17 of WKROUND 2009.

The estimates of high-grading by year are slightly revised when annual landings statistics are updated. In 2010, the time-series of estimates is:

| Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| HG $(\mathrm{t})$ | 210 | 148 | 74 | 432 | 592 | 322 | 25 |

In 2009, the low estimate of high-grading is likely to be related to the French vessels not being restricted by quota because of the decommissioning plan and the reports of effort directed towards more profitable species.

Both assumed Irish area misreporting and French high grading estimates since 2003 in percentages of the landings are summarized in the table below:

| Year | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\%$ | 3 | 7 | 4 | 17 | 30 | 29 | 2 |

High-grading also occurred in the UK catches in 2007-2008 but given the low level of landings, it has not been estimated.

The MLS of the Belgian landings is currently set at 40 cm .

## Fishery-science partnerships

## French self-sampling programme

In 2009, the French self-sampling program has been extended to several "métiers". The programme is voluntary under the auspices of the main Fishermen Organization P.M.A (Pêcheurs de Manche et Atlantique). In 2009, six otter trawlers have participated, providing data for métiers targeting either gadoids (OTB or OTTPD), Nephrops (OTTLN) or benthic species such as monkfish, megrim, rays, john dory (OTB or OTTPB). 38 trips were sampled in 2008 and 86 in 2009, summarized in the text table below. Because the sampling program is on voluntary basis, the métier targeting benthic species has been strongly sampled (43 trips by one trawler) though its contribution to the cod catches or landings is generally small.

| Gear Code | Q1 | Q2 |  | Q3 | Q4 | Total Métier |  |
| :--- | ---: | ---: | ---: | ---: | ---: | :--- | :---: |
| OTBPB | 7 | 15 | 14 | 7 | 43 | BENTH= OTBPB+OTTPB |  |
| OTBPD | 6 | 5 |  |  | 11 | GADI= OTBPD+OTTPD |  |
| OTTLN | 1 | 3 | 1 |  | 5 | NEPH= OTTLN |  |
| OTTPB | 1 |  | 3 | 2 | 6 |  |  |
| OTTPD | 8 | 6 | 5 | 4 | 23 |  |  |
| Total | 23 | 29 | 23 | 13 | 88 |  |  |

Several metiers can be fished during a single trip by changing fishing grounds (from fish to Nephrops for instance). Métiers have been identified by targeted species indicated by the skippers for each haul carried out.

2883 hauls have been sampled from 6022 carried out in the trips involved in the selfsampling programme. The sampling level for the Gadoid métier has fluctuated between 34 and $49 \%$ of hauls carried out. There is no sampling in the first quarter from the Nephrops trawlers because the methodology was more difficult and more time consuming to use in hauls where fish and Nephrops were always mixed. Results were better during the Nephrops season (Q2\&3) and poor in quarter 4 because of the heavy sea conditions. The number of hauls carried out and sampled is indicated in the text table below.

| Métier | Q1 | Q2 | Q3 | Q4 | Total |
| :--- | :---: | :---: | :---: | :---: | :---: |
| BENTH Total | 925 | 960 | 669 | 307 | 2861 |
| BENTH sampled | 231 | 559 | 501 | 266 | 1557 |
| GADI Total | 1147 | 1164 | 446 | 294 | 3051 |
| GADI sampled | 393 | 545 | 189 | 145 | 1272 |
| NEPH Total | 31 | 45 | 34 | $3^{*}$ | 110 |
| NEPH sampled | 0 | 29 | 24 | 1 | 54 |

three hauls targeting Nephrops in a GADI trip

Retained and discarded part of the catch have been scrutinized in each haul sampled. Overall 17215 cod have been measured, 15310 belonging to the retained part and 1905 to the discarded part.

In 2010, the self-sampling programme is continuing and the sampling data are input by the Professional Organization (P.M.A) in a database currently located at Ifre-
mer/Lorient. Motivation of the crew or the vessel owners could be a problem in future. The reasons are that 1) the effort of the industry to provide more biological data is not linked with incentives in setting TAC and quotas, 2) there has been in 2009 a pragmatic fit between the quota set and the fleet effort by change of métier or decommissioning which led to an under-consumption of the agreed quota, 3) the data collected by self-sampling have not been analyzed and used adequately because of the lack of official fisheries statistics data to raise properly the sampling data.

## Ireland-UK tagging programme in the Irish and Celtic Seas

The tagging programme focuses on both nursery areas and spawning aggregations of cod in the Irish and Celtic Seas, and involves conventional (plastic) tags and sophisticated electronic data storage tags. The programme in the Celtic Sea commenced in 2007 and is ongoing. The main objectives are to examine the movements of cod in relation to closed areas and in respect to stock mixing; to determine fine-scale movements and behaviour of cod during spawning; to examine vertical distribution (in relation to catchability) and thermal experiences (in relation to gonad development). Results of tagging work to date was presented to the ICES ASC in 2009 (Bendall et al., 2009). These results describe fundamental features of cod spatial ecology in the Irish and Celtic Sea, such as the location of feeding and spawning grounds (and the migratory pathways between them), the seasonality of migration and habitat occupation and the potential impact upon substock structure. Recaptures to date of juvenile cod tagged in the south of VIIa (Waterford estuary) shows that the majority of recaptures have occurred in VIIg mainly (O'Cuaig, Pers. comm.)

## Irish industry-science partnership quarter 1 cod survey

ICES (2009) notes that "given the uncertainty in the landings, the surveys represent the main source of information for estimating the historical trends in the stock." However, the current IBTS survey is conducted in quarter 4 when the stock is widely dispersed resulting in poor ability to track abundance due to low catch rates. ICES notes that "changing the surveys' design or programming additional stations are not thought to be relevant solutions, given the implications on other survey objectives" and ICES (2009) conclude that "adding a survey in quarter 1 would be the best solution, in order to monitor both the concentration of fish and the maturity during the spawning period." In recognition of this advice, the Marine Institute and the Federation of Irish Fishermen, in 2010 initiated an annual Q1 fishery independent survey for Celtic Sea Cod. The survey uses a commercial vessel and a dedicated survey trawl specification, based on a commercial design and in accordance with the criteria laid down in the ICES Study Group on Survey Trawl Standardisation (SGSST, 2009). The survey stations (Figure 7.2.1) are based on both Irish and foreign fleet VMS and/or logbook data. Using the VMS and logbook data, the Celtic Sea has been divided into areas of low, medium and high commercial catches and the survey sites have been randomly selected within these three categories (survey strata) with around $50 \%$ of the effort in the high areas and 30 and $20 \%$ in the medium and low (Figure 7.2.1). The data from the first survey is currently being worked up and will be presented to the WGCSE in 2011.

### 7.2.2 Data

## Landings

Tables 7.2.2 and 7.2.3 show the annual length structure of the landings per métier and country and the catch numbers-at-age respectively.

It is noticeable that this stock has always been composed of a few age classes. The catch number-at-age table (Table 7.2.3) shows the catch-was mainly composed of age 2 during the last 5 years. In 2009 the proportion of 2 year old fish is comparatively low and ages 3,4 , and 5 are higher than those observed since 2005.

## Discards

Table 7.2.4 and Figure 7.2.2a-c show the length structure of landings and discards per country and quarter with a split by métier for France (No 2009 update for France was available at the time of the meeting). French information is split into self-sampling (Figure 7.2.2d) and on-board observer programmes, noting that the latter is incomplete, because validation of some trips is still ongoing. It is noticeable that the majority of the cod discarded result from the high-grading behaviour, for France and UK. Discarding of undersized individuals is at low level for all countries.

## Biological

Catch in numbers-at-age (Figure 7.2.3) and stock weights are given respectively in Tables 7.2.5, 7.2.6. The final year estimates are consistent with the recent historical values.

Natural mortality, percentage of F before spawning and maturity ogive remained unchanged and are described in the Stock Annex. Celtic Sea cod are very fast growing and early maturing compared with more northern cod stocks.

## Surveys

Tables 7.2.7 present the survey dataseries.
Internal consistency of the two ongoing surveys (FR-EVHOE \& IR-GFS7gj combined) has been explored using SURBA software. The number of fish sampled during those surveys remains low as those species are not specifically targeted.

The raw abundance indices (number of individuals caught per 30 minutes tow) of FREVHOE have been provided to the WG. Indices have an average CV of around $25 \%$ and have changed since 2002 within the confidence intervals (bottom right of Figure 7.2 .4 a ). CV were calculated taking account the surface of each stratum, total surface, sample mean catch per tow, variance of the catch and number of tow for each stratum according to the method in Cochran (1979).

Figure 7.2.4a summarises the single fleet analysis for FR-EVHOE. The tracking of recruitment is well defined for the relatively good YC 1996, 1999 and 2000, and poor YC 2001 and 2002, especially at age 1 . The weakness seems to be in-between year consistency especially for the older ages. The log residuals show a low level of noise, resulting from the recurrent low catch rates.

Figure 7.2.4b represents the single fleet analysis for IR-GFS7gj. The short time-series prevents conclusions on the consistency, but the tracking of recent year classes is consistent with FR-EVHOE except for the 2007 YC.

The former UK-WCGFS was also included in the analysis to smooth the signal when looking at historical trends in the stock (Figure 7.2.4c).

Figure 7.2.4d represents SURBA model estimates of mean Z for the three single fleets. Each time-series of $Z$ fluctuates within the magnitude of the uncertainty, resulting in non-robust general trends. Moreover, SURBA is known to provide poor estimates of
parameters for the most recent years. As a result, no clear trend can be seen from the surveys.

Figure 7.2.4e shows the comparative analysis of the age 1 index from the FR-EVHOE survey and the estimates of recruitment from a Separable VPA summary. FR-EVHOE has demonstrated some ability to predict the level of expected recruitment but the recruitment indices have been diverging in recent years reflecting higher rates of discarding and high-grading in the catch data.

Overall, no clear trend of change in biomass or mortality can be derived from any of the survey indices.

## Commercial cpue

Tables 7.2.8a, b and c show the series of landings, fishing effort and lpue dataseries for four French fleets, three UK fleets and eight Irish fleets. Figure 7.2.5a and b show their trends. French catch and effort data for 2009 were not available at the time of the meeting. A general decrease in the lpue trend is observed in almost all series between 1990 and 2004, where the TAC began to be constraining. From that point, the lpues seemed to stabilize, or even to increase if high grading is taken into account.

Different features are observed in the effort time-series. The métiers showing the highest levels of cod directed effort have decreased significantly in the last 5-10 years. Irish otter shows an increasing trend over the period, the majority of this effort is directed towards Nephrops.

A special effort has been made during the 2009 WG to combine international landings and effort datasets and produce historical distribution maps. These maps are respectively composed of France, UK, Ireland and Belgium landings (Figure 7.2.6), France and Ireland efforts (Figure 7.2.7) and lpue (Figure 7.2.8). The data are not corrected for misreporting or high-grading. The main outcome of these maps is the shrinking of the geographical area of the stock over the years. This is particularly visible in the distribution of the landings (Figure 7.2.6). The perceived decrease of landings over time is to be regarded with caution given the recent levels of misreporting and highgrading. The rectangles temporarily closed (30E4, 31E4 and 32E3) since 2005 were clearly among the most important in terms of lpue.


Green: Trevose closed areas.

### 7.2.3 Stock assessment

Model used: None.

No analytical assessment has been carried out on this stock, following the recommendations from WKROUND 2009 and the lack of revision of the available datasets.

## Exploratory analysis on the MSY approach

Following the guidelines defined during WKFRAME, some exploratory work has been done on this stock in an attempt to define MSY indicators. The main issue for this stock is the lack of assessment due to the quality of recent data. Therefore, the range of methods is limited and this analysis can only be classified as exploratory.

As suggested by WKFRAME for stocks with age and length, YPR has been used to explore the expected yield under equilibrium conditions, of growth, maturity and natural mortality, for a given or assumed fishery pattern, across a range of exploitation levels.

YPR (http://nft.nefsc.noaa.gov/) requires age-structured data which were provided from the updated dataset that are used for the exploratory VPA (Table 7.2.9). Data were checked against the trends in the last ten years as well as over the full timeseries. The age structure for cod has not experienced substantial changes through time except in abundance. Analyses from both 10 and 38 years long time-series (19912009) show a total mortality respectively between 1.05 and 1.02 (Figure 7.2.9).

With a natural mortality of $\mathrm{M}=0.2$, assuming a fishing mortality between 0.82 and 0.85 , the fishing mortality appears twice above the value of $F_{\max }$ and three times the value of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{40 \%}$.

|  | F | Y/R | SSB/R | Tot. Biom/R | Mean age | Mean gener. | Expected |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | time | spawning |
| F $_{0}$ | 0 | 0 | 72.15 | 76.81 | 5.17 | 7.11 | 6.40 |
| F-01 | 0.27 | 4.19 | 29.75 | 33.29 | 3.98 | 6.48 | 2.66 |
| F Max | 0.38 | 4.35 | 21.63 | 24.90 | 3.58 | 6.22 | 1.95 |
| F at 40\% | 0.28 | 4.22 | 28.86 | 32.38 | 3.94 | 6.46 | 2.58 |
| MSP |  |  |  |  |  |  |  |

The $F_{\text {max }}$ peak appears well defined (Figure 7.2.10) and running the same analysis for both 10 and 38 years time-series does not substantially change results. This YPR analysis also falls into the category of stocks where discards are substantial. A sensitivity analysis to the natural mortality has been performed with $M$ varying between 0.1 and 0.3 . Values of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ are not substantially affected by the change of Mvalue (Figure 7.2.11).

These $\mathrm{F}_{\max }$ and $\mathrm{F}_{0.1}$ estimates are very much in line with those obtained from historical XSA assessments for this stock (see previous WGSSDS reports). Fishing mortality from historical XSA assessments have been well above $\mathrm{F}_{\text {max }}$. and in line with those obtained here.

### 7.2.4 Short-term projections

No short-term projections were carried out.

### 7.2.5 Medium-term projection

No medium-term projections were carried out.

### 7.2.6 Biological reference points

WKROUND 2009 has suggested that, unless there is an investigation on the possible change in the maturity ogive, there was no solid reason to change the biological reference points. The biological reference points are then recalled below:

| Ref. point | ACFM 1998 | WG 1999* | ACFM 1999 | WG 2004 | ACFM 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Flim | $0.90 \text { (Floss }$ WG98) | 0.90 (history <br> WG99) | $0.90 \text { (history }$ WG99) |  | 0.90 (history WG99) |
| $\mathrm{F}_{\mathrm{pa}}$ | 0.68 (5th perc <br> Floss WG98) | 0.65 (Flim*0.72) | 0.68 (5th perc <br> Floss WG98) |  | 0.68 (5th perc Floss WG98) |
| Blim | $\begin{aligned} & 4500 \mathrm{t} \text { (Bloss } \\ & =\text { B76 WG98) } \end{aligned}$ | $\begin{aligned} & 5400 \mathrm{t} \text { (Bloss=B76 } \\ & \text { WG99) } \end{aligned}$ | $\begin{aligned} & 5400 \mathrm{t} \\ & \text { (Bloss=B76 } \\ & \text { WG99) } \end{aligned}$ | 6300 t <br> (Bloss=B76 <br> WG04) | 6300 t <br> (Bloss=B76 <br> WG04) |
| $\mathrm{B}_{\text {pa }}$ | $\begin{aligned} & 8000 \mathrm{t} \\ & \left(\operatorname{Blim}^{*} 1.65\right) \end{aligned}$ | $\begin{aligned} & 9000 \mathrm{t} \\ & \left(\text { Blim }^{*} 1.65\right) \end{aligned}$ | 10000 t <br> (history) | Reject - no SR relation | $\begin{aligned} & 8800 \mathrm{t}(\mathrm{Bpa}= \\ & \left.\mathrm{Blim}^{*} 1.4\right) \end{aligned}$ |

### 7.2.7 Management plans

A long-term management plan has been under discussion for this stock and an effort based management system in the Celtic Sea (VIIfg) is being discussed by member states and the EC.

### 7.2.8 Uncertainties and bias in assessment and forecast

The assessment of this stock is impaired by a strong uncertainty in the level of catches, especially since the TAC became constraining from 2003 onward. For this reason, and until a more reliable information is available, WKROUND 2009 concluded that the current assessment procedure treating catch numbers as unbiased was no longer appropriate. Surveys lack robust trends mainly due to their low catch rates.

### 7.2.9 Recommendation for next Benchmark

This stock should be benchmarked with the other WGCSE cod stocks in late 2011 or 2012.

|  |  |  |  |
| :--- | :--- | :--- | :--- |
| Cod | WKROUND 2009 concluded that more work is required <br> before Celtic Sea cod can be benchmarked successfully. <br> WGCSE 2010 reviewed the available infromation and several <br> improvements have occurred since WKROUND. First there is <br> now a time-series of self-sampling high-grading estimates. <br> Discard and misreporting rates appear to have declined. has <br> suggested a response to their recommendation in Section 7.2. | Late | Expert <br> Group <br> members |
|  | There is a growing body of new tagging information that may <br> prove useful to assess stock structure and possible mortality <br> rate. | There is a new dedicated survey for the stock that need to be <br> considered and the two other IBTS survey-series should be <br> examined to see if a comboned index might be possible. | experts |

### 7.2.10 Management considerations

Fishing mortality from historical assessments have been well above potential $\mathrm{F}_{\text {msy }}$ proxies for this stock. It is not possible to determine current fishing mortality rates due uncertain catch-at-age data and surveys. This was also the case last year when ICES advised "that fishing effort and catches should be reduced although it is not possible to determine the appropriate scale of such reduction".

The geographical range of the stock appears to have contracted significantly according to the international landings and lpue distribution maps. This stock has had a very truncated age structure with age 2 fish being the most numerous in landings over many years. The historical dynamics of Celtic Sea cod have been "recruitment driven", i.e. the stock increased in the past in response to good recruitments and decreased rapidly during times of poor recruitment. Recruitment in recent years appears to be poor. Fishing mortality should be reduced in the longer term to maximize the contributions of recruitment to future SSB and yield and will result in reduced risk to the stock.

Cod in Divisions VIIe- k are caught in a range of fisheries including gadoid trawlers, Nephrops trawlers, otter trawlers, beam trawlers, and gillnetters. Other commercial species that are caught by these fisheries include haddock, whiting, Nephrops, plaice, sole, anglerfish, hake, megrim, and elasmobranchs.

In the recent past there have been indications of an underreporting of cod landings in some fleets. The introduction of the buyers and sellers legislation in the UK and Ireland may have reduced this, but may also have increased discards. Measures aimed at reducing discarding and improving the fishing pattern should be encouraged. These might include spatial and temporal changes in fishing practices or technical measures. These measures would need to be evaluated in the context of other species caught in mixed fisheries.

The exclusion of ICES Division VIId in the TAC area since 2009 makes the management area more in line with the boundaries of the stock as the stock is VIId is considered as an extension of the cod population in the North Sea.

Since 2005, ICES rectangles 30E4, 31E4, and 32E3 have been closed during the first quarter (Council Regulations 27/2005, 51/2006, and 41/2007, 40/2008 and 43/2009) with the objective of reducing fishing mortality on cod. At an annual resolution, maps of international effort distribution do not show evidence that this closure has redistributed effort of otter trawlers to other areas.

There have been major changes in fleet dynamics over the period of the assessment. Effort in the French otter trawlers has been declining since 1999 and a decommissioning plan has occurred in 2008 and a new plan is ongoing since 2009. A consequence of the Trevose closure is that a part of the effort displayed by the French otter trawlers in the three rectangles before or after the closure has been reported to the allowed area where the catch of mixed species (mainly gadoids) is still profitable, particularly in the rectangles neighbouring the closed area (rectangles $32 \mathrm{E} 4,32 \mathrm{E} 2,31 \mathrm{E} 2,31 \mathrm{E} 3$, $30 \mathrm{E} 3,29 \mathrm{E} 3,29 \mathrm{E} 4$ ) or in a more distant and still shallower rectangle 31E1. Another part of the effort is displayed in the rectangles 29E1, 28E1, meaning that this effort is then targeting Nephrops, monkfish, megrim, Nephrops and elasmobranch. Overall, a part of the French bottom trawlers has not changed their activity with the closed period and continue to target gadoid fish in the neighbouring rectangles of the closed area. Another part of them target benthic species (anglerfish, megrim and john dory) in more distant rectangles 28E1, 29E1.

Irish otter trawl effort in VIIg,j has been stable over the last four years. During this period there has been a fleet modernisation and several decommissioning schemes in Ireland both within the national whitefish fleet and beam trawl fleet.

### 7.2.11 References

Bendall, V., O Cuaig, M, Schön, P-J., Hetherington, S., Armstrong, M., Graham, N., and Righton, D. 2009. Spatiotemporal dynamics of Atlantic cod (Gadus morhua) in the Irish and Celtic Seas: results from a collaborative tagging programme ICES CM 2009/J:06

Cochran, W.G. 1977. Sampling Technics. J. Wiley \& Sons. 428 p.

Table 7.2.1. Nominal landings of Cod in Divisions VII e-k used by the Working Group.

| 5782 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 4737 |  |  |  |  |  |  |  |  |
| 1973 | 4015 |  |  |  |  |  |  |  |  |
| 2898 |  |  |  |  |  |  |  |  |  |
| 1975 | 3993 |  |  |  |  |  |  |  |  |
| 1976 | 4818 |  |  |  |  |  |  |  |  |
| 1977 | 3058 |  |  |  |  |  |  |  |  |
| 1978 | 3647 |  |  |  |  |  |  |  |  |
| 1979 | 4650 |  |  |  |  |  |  |  |  |
| 7243 |  |  |  |  |  |  |  |  |  |
| 10596 |  |  |  |  |  |  |  |  |  |
| 1982 | 8766 |  |  |  |  |  |  |  |  |
| 9641 |  |  |  |  |  |  |  |  |  |
| 1984 | 6631 |  |  |  |  |  |  |  |  |
| 1985 | 8317 |  |  |  |  |  |  |  |  |
| 10475 |  |  |  |  |  |  |  |  |  |
| 10228 |  |  |  |  |  |  |  |  |  |
| 1988554 | 13863 | 1480 | 12922 | 17191 |  |  |  |  |  |
| 1989910 | 15801 | 1860 | 122315 | 19809 |  |  |  |  |  |
| 1990621 | 9383 | 1241 | 1346158 | 12749 |  |  |  |  |  |
| 1991303 | 6260 | 1659 | 109420 | 9336 |  |  |  |  |  |
| 1992195 | 7120 | 1212 | 120713 | 9747 |  |  |  |  |  |
| 1993391 | 8317 | 766 | 9456 | 10425 |  |  |  |  |  |
| 1994398 | 7692 | 1616 | 9068 | 10620 |  |  |  |  |  |
| 1995400 | 8321 | 1946 | 10348 | 11709 |  |  |  |  |  |
| 1996552 | 8981 | 1982 | 11660 | 12680 |  |  |  |  |  |
| 1997694 | 8662 | 1513 | 11660 | 12035 |  |  |  |  |  |
| 1998528 | 8096 | 1718 | 10890 | 11431 |  | Benchmark 2009 WGCSE |  |  |  |
| 1999326 | 5488 | 1883 | 8970 | 8594 |  | HG based | 2008 | HG based | 2009 |
| 2000208 | 4281 | 1302 | 7440 | 6535 |  | self sampling data 2009 FR self |  |  |  |
| 2001347 | 6033 | 1091 | 8380 | 8309 HG based on UK |  |  |  |  |  |
| 2002555 | 7368 | 694 | 6180 | 9235 Highgrading Total Highgrading Total Highgrading Total |  |  |  |  |  |
| 2003136 | 5222 | 517 | 3460 | 62212106431 |  |  |  |  |  |
| 2004153 | 2425 | 663 | 2820 | 3523 | 148 | 3671 |  |  |  |
| 2005186 | 1623 | 870 | 3090 | 2988 | 74 | 3062 |  |  |  |
| 2006103 | 1896 | 959 | 3680 | 3326 |  | 4323758 |  |  |  |
| 2007108 | 2509 | 1210 | 4120 | 4239 |  | 5924831 |  |  |  |
| 200865 | 2064 | 1221 | 2890 | 3639 |  | 322 396 |  |  |  |
| 2009* 49 | 2027 | 870 | 2640 | 3210 |  |  |  | 25 | 3235 |

* provisional

Updated for WKROUND and WGCSE 2009
Scaled landings 1971-1987 (SSDS WG 1999)

Table 7.2.2. Cod in Divisions VIIe-k. 2009 Landings in numbers-at-length (cm). Note: French data by metier were not available at the time of the meeting. é

|  | France | UK VII e-k | UK VII e-k | Ireland |
| :---: | :---: | :---: | :---: | :---: |
|  | VII e-k | Beam trawl | All bar beam trawl | VIIg,j |
| Length (cm) |  |  |  |  |
| 24 |  |  |  |  |
| 25 |  |  |  |  |
| 26 |  |  |  |  |
| 27 |  |  |  |  |
| 28 |  |  |  |  |
| 29 |  |  |  | 0 |
| 30 |  |  |  | 0 |
| 31 | 18 |  | - | 42 |
| 32 | 18 |  | - | 55 |
| 33 | 0 |  |  | 51 |
| 34 | 0 |  | 17 | 289 |
| 35 | 349 | 8 | 5 | 1154 |
| 36 | 1429 | 24 | 45 | 1801 |
| 37 | 2087 | 64 | 204 | 2589 |
| 38 | 4553 | 126 | 298 | 5046 |
| 39 | 4249 | 181 | 225 | 4710 |
| 40 | 5573 | 149 | 353 | 7053 |
| 41 | 5117 | 222 | 437 | 6644 |
| 42 | 5711 | 339 | 360 | 5904 |
| 43 | 4484 | 389 | 582 | 5302 |
| 44 | 4223 | 414 | 634 | 6591 |
| 45 | 5281 | 310 | 346 | 7133 |
| 46 | 5501 | 365 | 293 | 6373 |
| 47 | 5515 | 376 | 268 | 4133 |
| 48 | 5280 | 218 | 710 | 4608 |
| 49 | 4125 | 347 | 428 | 3623 |
| 50 | 5847 | 217 | 629 | 3832 |
| 51 | 5511 | 382 | 566 | 3388 |
| 52 | 6396 | 386 | 741 | 3115 |
| 53 | 8225 | 406 | 743 | 3585 |
| 54 | 8747 | 419 | 1193 | 2668 |
| 55 | 10836 | 383 | 938 | 3153 |
| 56 | 11693 | 488 | 1154 | 3569 |
| 57 | 10527 | 436 | 1283 | 3475 |
| 58 | 14229 | 418 | 974 | 3524 |
| 59 | 15250 | 583 | 1028 | 4023 |
| 60 | 17575 | 662 | 878 | 3614 |
| 61 | 16721 | 588 | 1229 | 4138 |
| 62 | 16312 | 523 | 1051 | 4279 |
| 63 | 14190 | 478 | 1077 | 4483 |
| 64 | 14602 | 539 | 873 | 6065 |


|  | France | UK VII e-k | UK VII e-k | Ireland |
| :---: | :---: | :---: | :---: | :--- |
|  | VII e-k | Beam trawl | All bar beam trawl | VIIg,j |
| Length (cm) |  |  |  |  |
| 65 | 13657 | 551 | 1035 | 4471 |
| 66 | 13095 | 489 | 1290 | 5844 |
| 67 | 12916 | 481 | 949 | 5308 |
| 68 | 13111 | 417 | 874 | 5120 |
| 69 | 12811 | 562 | 1111 | 4576 |
| 70 | 13704 | 409 | 1292 | 5574 |
| 71 | 11205 | 397 | 1326 | 5418 |
| 72 | 12923 | 395 | 1419 | 5253 |
| 73 | 12264 | 315 | 1470 | 5069 |
| 74 | 11517 | 314 | 1653 | 6059 |
| 75 | 12242 | 236 | 1411 | 6281 |

Table 7.2.2. Continued.

|  | France | UK VII e-k | UK VII e-k | Ireland |
| :---: | :---: | :---: | :---: | :---: |
|  | VII e-k | Beam trawl | All bar beam trawl | VIIg,j |
| Length (cm) |  |  |  |  |
| 76 | 12600 | 266 | 1272 | 4502 |
| 77 | 9595 | 301 | 1393 | 4000 |
| 78 | 11318 | 333 | 958 | 3760 |
| 79 | 8562 | 279 | 1319 | 2723 |
| 80 | 11945 | 438 | 1023 | 2916 |
| 81 | 6698 | 343 | 916 | 2957 |
| 82 | 7811 | 266 | 530 | 2699 |
| 83 | 7776 | 177 | 620 | 2582 |
| 84 | 7043 | 196 | 630 | 2431 |
| 85 | 6357 | 152 | 688 | 2418 |
| 86 | 6317 | 168 | 504 | 3070 |
| 87 | 5344 | 138 | 368 | 2804 |
| 88 | 6479 | 122 | 444 | 2589 |
| 89 | 5229 | 101 | 293 | 2619 |
| 90 | 5907 | 73 | 366 | 1503 |
| 91 | 3481 | 88 | 241 | 1453 |
| 92 | 3643 | 113 | 214 | 1857 |
| 93 | 3243 | 116 | 142 | 1395 |
| 94 | 3046 | 55 | 200 | 1558 |
| 95 | 2600 | 99 | 178 | 1208 |
| 96 | 1894 | 61 | 148 | 416 |
| 97 | 2206 | 53 | 181 | 632 |
| 98 | 1900 | 51 | 111 | 958 |
| 99 | 1869 | 66 | 68 | 298 |
| 100 | 2536 | 41 | 65 | 1040 |
| 101 | 939 | 48 | 44 | 394 |
| 102 | 1001 | 11 | 37 | 257 |


|  | France | UK VII e-k | UK VII e-k | Ireland |
| :---: | :---: | :---: | :---: | :---: |
|  | VII e-k | Beam trawl | All bar beam trawl | VIIg,j |
| Length (cm) |  |  |  |  |
| 103 | 808 | 17 | 69 | 301 |
| 104 | 726 | 28 | 23 | 262 |
| 105 | 661 | 23 | 0 | 198 |
| 106 | 226 | 22 | 9 | 104 |
| 107 | 267 | 14 | 0 | 109 |
| 108 | 48 | 11 | 0 | 93 |
| 109 | 13 | 11 | 14 | 0 |
| 110 | 105 | 20 |  | 83 |
| 111 | 65 |  |  |  |
| 112 | 92 |  |  |  |
| 113 | 54 |  |  |  |
| 114 | 0 |  |  |  |
| 115 | 30 |  |  |  |
| 116 | 0 |  |  |  |
| 117 | 55 |  |  |  |
| 118 | 50 |  |  |  |
| 119 | 0 |  |  |  |
| 120 | 9 |  |  |  |
| 121 | 0 |  |  |  |
| 122 | 22 |  |  |  |
| 123 | 0 |  |  |  |
| 124 | 0 |  |  |  |
| 125 | 28 |  |  |  |
| Total | 530215 | 19307 | 46460 | 245174 |
| Tw | 2027.1 | 67.3 | 178 | 653.8 |
| Mean length | 67.6 | 64.5 | 67.5 | 63.0 |
| Mean Weight | 3.823 | 3.486 | 3.831 | 2.667 |

Table 7.2.3. Cod in Divisions VIIe-k (Celtic Sea). Catch numbers-at-age. Area reallocation (IRL 2004 to 2009) and high-grading (FR 2003-2009) included.

Run title : Cod in Divisions VIIe-k,WGCSE10, index file

At 6/05/2010 12:02

| Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1971, | 1972, | 1973, | 1974, | 1975, | 1976, | 1977, | 1978, | 1979, |
| AGE |  |  |  |  |  |  |  |  |  |
| 1, | 725, | 4, | 332, | 1, | 673, | 51, | 25, | 197, | 438, |
| 2, | 461, | 774, | 239, | 224, | 136, | 1456, | 416, | 497, | 357 |
| 3, | 557, | 110, | 346, | 40, | 185, | 61, | 236, | 129, | 263 |
| 4, | 96, | 205, | 60, | 118, | 61, | 107, | 15, | 116, | 68 |
| 5, | 35, | 45, | 74, | 38, | 105, | 11, | 60 , | 20, | 104 |
| 6, | 17, | 26, | 17, | 37, | 20, | 22, | 2, | 34, | 19 |
| +gp, | 11, | 17, | 11, | 36, | 33, | 7, | 17, | 20, | 32 |
| TOTALNUM, | 1902, | 1181, | 1079, | 494, | 1213, | 1715, | 771, | 1013, | 1281 |
| TONSLAND, | 5782, | 4737, | 4015, | 2898, | 3993, | 4818, | 3059, | 3647, | 4650 |
| SOPCOF \%, | 100, | 100, | 100, | 100, | 100, | 100, | 100, | 100, | 100 |

Table 1 Catch numbers at age Numbers*10**-3
YEAR, 1980, 1981, 1982, 1983, 1984, 1985, 1986, 1987, 1988, 1989,


## Run title : Cod in Divisions VIIe-k,WGCSE10,index file

```
At 6/05/2010 12:02
```

    Table 1 Catch numbers at age Numbers*10**-3
    YEAR, 1990, 1991, 1992, 1993, 1994, 1995, 1996, 1997, 1998, 1999,
    AGE
    1, 360, 1377, 1434, 274, 1340, 823, 617, 1184, 639, 496,
    

Table 7.2.4a. Cod in Divisions VIIe-k. Length structure of landings and discards from sampling by UK.

UK - Sampled data raised to trips sampled.

| Length | Q1 |  | Q2 |  | Q3 |  | Q4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (cm) | Retained | Discarded | Retained | Discarded | Retained | Discarded | Retained | Discarded |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 1 |
| 14 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| 16 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 |
| 17 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| 23 | 0 | 1 | 0 | 2 | 0 | 1 | 0 | 0 |
| 24 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 25 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
| 26 | 0 | 4 | 0 | 3 | 0 | 0 | 0 | 0 |
| 27 | 0 | 2 | 2 | 2 | 0 | 0 | 0 | 0 |
| 28 | 0 | 2 | 2 | 2 | 0 | 0 | 0 | 0 |
| 29 | 0 | 7 | 0 | 1 | 0 | 1 | 0 | 0 |
| 30 | 0 | 2 | 0 | 1 | 0 | 1 | 0 | 0 |
| 31 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 |
| 32 | 0 | 1 | 0 | 0 | 0 | 7 | 0 | 0 |
| 33 | 0 | 0 | 0 | 2 | 0 | 7 | 0 | 0 |
| 34 | 0 | 0 | 0 | 4 | 1 | 0 | 0 | 0 |
| 35 | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 4 |
| 36 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 37 | 1 | 0 | 0 | 0 | 6 | 3 | 1 | 0 |
| 38 | 1 | 0 | 0 | 0 | 0 | 6 | 0 | 0 |
| 39 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |


| Length | Q1 |  | Q2 |  | Q3 |  | Q4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (cm) | Retained | Discarded | Retained | Discarded | Retained | Discarded | Retained | Discarded |
| 40 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 |
| 41 | 0 | 1 | 1 | 0 | 5 | 0 | 4 | 0 |
| 42 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |
| 44 | 1 | 6 | 0 | 0 | 1 | 0 | 1 | 0 |
| 45 | 0 | 0 | 1 | 0 | 2 | 1 | 0 | 0 |
| 46 | 3 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |
| 47 | 2 | 1 | 0 | 0 | 0 | 0 | 4 | 0 |
| 48 | 0 | 5 | 0 | 0 | 0 | 1 | 0 | 0 |
| 49 | 0 | 11 | 1 | 0 | 0 | 3 | 0 | 0 |
| 50 | 3 | 12 | 0 | 0 | 1 | 2 | 0 | 0 |
| 51 | 2 | 20 | 1 | 0 | 1 | 5 | 0 | 0 |
| 52 | 1 | 24 | 2 | 0 | 0 | 5 | 0 | 0 |
| 53 | 5 | 16 | 2 | 0 | 0 | 28 | 0 | 0 |
| 54 | 4 | 12 | 4 | 0 | 0 | 9 | 1 | 0 |
| 55 | 1 | 17 | 7 | 1 | 3 | 7 | 0 | 0 |
| 56 | 4 | 9 | 2 | 1 | 4 | 7 | 0 | 0 |
| 57 | 5 | 7 | 6 | 0 | 3 | 7 | 0 | 0 |
| 58 | 3 | 4 | 9 | 0 | 2 | 8 | 0 | 0 |
| 59 | 2 | 1 | 5 | 0 | 1 | 11 | 0 | 0 |
| 60 | 5 | 1 | 3 | 0 | 4 | 5 | 0 | 0 |
| 61 | 5 | 0 | 1 | 0 | 2 | 6 | 0 | 0 |
| 62 | 6 | 0 | 5 | 0 | 1 | 7 | 0 | 0 |
| 63 | 5 | 0 | 1 | 0 | 2 | 2 | 0 | 0 |
| 64 | 5 | 0 | 1 | 0 | 4 | 8 | 0 | 0 |
| 65 | 4 | 0 | 3 | 0 | 0 | 1 | 0 | 0 |
| 66 | 9 | 0 | 1 | 0 | 1 | 0 | 4 | 0 |
| 67 | 6 | 0 | 2 | 0 | 5 | 0 | 0 | 0 |
| 68 | 13 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| 69 | 9 | 1 | 1 | 0 | 2 | 1 | 0 | 0 |
| 70 | 12 | 0 | 0 | 1 | 2 | 0 | 0 | 0 |
| 71 | 9 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 72 | 11 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
| 73 | 7 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 74 | 6 | 0 | 1 | 1 | 0 | 0 | 1 | 0 |
| 75 | 13 | 0 | 5 | 1 | 1 | 0 | 0 | 0 |

Table 7.2.4a. Continued.

UK - Cod VIIe-k - Sampled data raised to trips sampled.

| Length | Q1 |  | Q2 |  | Q3 |  | Q4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (cm) | Retained | Discarded | Retained | Discarded | Retained | Discarded | Retained | Discarded |
| 76 | 6 | 0 | 1 | 1 | 1 | 0 | 0 | 0 |
| 77 | 12 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 78 | 5 | 0 | 3 | 0 | 1 | 0 | 0 | 0 |
| 79 | 2 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |
| 80 | 2 | 0 | 1 | 1 | 1 | 0 | 0 | 0 |
| 81 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 82 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 83 | 3 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 84 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
| 85 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 86 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| 87 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 88 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 89 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 90 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 91 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 92 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 93 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 94 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 95 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 96 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 97 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 98 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 99 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 100 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
| 101 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 102 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 103 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 104 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 105 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 106 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 108 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 109 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 110 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 111 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 112 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 113 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Total N | 221 | 172 | 82 | 24 | 76 | 166 | 18 | 6 |
| Trips | 32 |  | 22 |  | 29 |  | 12 |  |

Table 7.2.4b. Cod in Divisions VIIe-k. Length structure of landings and discards from sampling by Ireland.

## Cod in ICES Division VIIg

Irish Otter Trawl Discard Numbers and Mean Weights-at-Age, and Raised length distribution
No of Trips= Sampled 16 Logbooks 2641
No. of hauls $=175$

## Fishing Year 2009

Total Otter Trawl Discards $=82$ tonnes [live]
Otter Trawl, Irish vessels landing into Irish and Foreign Ports
Raised (using Trip as Variable)

|  | Frequency ('000) |  |  | Frequency ('000) |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Raised Discards | Retained Catch | Length (cm) | Raised Discards | Retained Catch |
| 25 | 4.74 | 0.00 | 66 | 0.00 | 1.55 |
| 26 | 2.98 | 0.00 | 67 | 0.00 | 0.64 |
| 27 | 4.27 | 0.00 | 68 | 0.00 | 1.17 |
| 28 | 7.43 | 0.00 | 69 | 0.00 | 0.58 |
| 29 | 0.68 | 0.00 | 70 | 0.00 | 1.79 |
| 30 | 3.09 | 0.92 | 71 | 0.00 | 0.90 |
| 31 | 12.17 | 0.00 | 72 | 0.00 | 0.49 |
| 32 | 15.69 | 0.00 | 73 | 0.00 | 0.26 |
| 33 | 1.74 | 0.00 | 74 | 0.00 | 0.36 |
| 34 | 2.60 | 0.00 | 75 | 0.00 | 0.87 |
| 35 | 7.05 | 0.92 | 76 | 0.00 | 1.22 |
| 36 | 16.54 | 0.92 | 77 | 0.00 | 1.02 |
| 37 | 33.03 | 0.00 | 78 | 0.00 | 0.69 |
| 38 | 9.82 | 0.04 | 79 | 0.00 | 1.41 |
| 39 | 23.68 | 0.07 | 80 | 0.00 | 0.36 |
| 40 | 4.21 | 0.06 | 81 | 0.00 | 1.19 |
| 41 | 2.09 | 0.11 | 82 | 0.00 | 0.58 |
| 42 | 25.80 | 0.45 | 83 | 0.00 | 0.33 |
| 43 | 3.20 | 0.31 | 84 | 0.00 | 0.51 |
| 44 | 25.19 | 0.71 | 85 | 0.00 | 0.30 |
| 45 | 0.00 | 0.66 | 86 | 0.00 | 0.25 |
| 46 | 0.00 | 0.55 | 87 | 0.00 | 1.17 |
| 47 | 0.00 | 0.27 | 88 | 0.00 | 0.31 |
| 48 | 0.00 | 0.51 | 89 | 0.00 | 0.34 |
| 49 | 0.00 | 0.89 | 90 | 0.00 | 0.48 |
| 50 | 0.00 | 0.38 | 91 | 0.00 | 0.34 |
| 51 | 0.00 | 0.29 | 92 | 0.00 | 0.36 |
| 52 | 0.00 | 0.44 | 93 | 0.00 | 0.76 |
| 53 | 0.00 | 0.74 | 94 | 0.00 | 0.34 |
| 54 | 0.00 | 0.59 | 95 | 0.00 | 0.49 |
| 55 | 0.00 | 0.90 | 96 | 0.00 | 0.28 |
| 56 | 0.00 | 0.13 | 97 | 0.00 | 0.32 |
| 57 | 0.00 | 0.67 | 98 | 0.00 | 0.08 |


|  | Frequency ('000) |  | Frequency ('000) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | Raised Discards Retained Catch | Length (cm) | Raised Discards Retained Catch |  |  |
| 58 | 0.00 | 0.40 | 99 | 0.00 | 0.28 |
| 59 | 0.00 | 0.32 | 100 | 0.00 | 0.20 |
| 60 | 0.00 | 0.90 | 101 | 0.00 | 0.34 |
| 61 | 0.00 | 0.76 | 102 | 0.00 | 0.44 |
| 62 | 0.00 | 0.62 | Total | 205.99 | 38.86 |
| 63 | 0.00 | 0.28 |  |  |  |
| 64 | 0.00 | 0.43 |  |  |  |
| 65 | 0.00 | 0.62 |  |  |  |

Table 7.2.4c. Cod in Divisions VIIe-k. Length structure of French landings and discards from the self-sampling Program. Sampling data raised by landing ratio to the total catch of the fleet in VIIfgh.

| 2008 | Retained |  |  |  | Discarded |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | FR-GADOID |  |  |  | FR-GADOID |  |  |  |  |
|  | 2008- Q1 | 2008 - Q2 | 2008- Q3 | 2008-Q4 |  | 2008-Q1 | 2008 - Q2 | 2008-Q3 | 2008 - Q4 |
| Length <br> (cm) |  |  |  |  | Lengt (cm) |  |  |  |  |
| 20 |  |  |  |  | 20 | 0 | 0 | 0 | 0 |
| 21 |  |  |  |  | 21 | 0 | 0 | 0 | 0 |
| 22 |  |  |  |  | 22 | 0 | 0 | 0 | 0 |
| 23 |  |  |  |  | 23 | 0 | 0 | 0 | 0 |
| 24 |  |  |  |  | 24 | 0 | 40 | 0 | 0 |
| 25 |  |  |  |  | 25 | 0 | 20 | 0 | 0 |
| 26 |  |  |  |  | 26 | 0 | 0 | 0 | 0 |
| 27 |  |  |  |  | 27 | 46 | 20 | 14 | 0 |
| 28 |  |  |  |  | 28 | 0 | 60 | 41 | 15 |
| 29 |  |  |  |  | 29 | 69 | 40 | 0 | 107 |
| 30 |  |  |  |  | 30 | 161 | 40 | 96 | 168 |
| 31 |  |  |  |  | 31 | 323 | 241 | 329 | 259 |
| 32 |  |  |  |  | 32 | 576 | 181 | 233 | 229 |
| 33 |  |  |  |  | 33 | 1060 | 140 | 575 | 305 |
| 34 |  |  |  |  | 34 | 1129 | 281 | 726 | 442 |
| 35 |  |  |  |  | 35 | 1498 | 381 | 849 | 458 |
| 36 |  |  |  |  | 36 | 2235 | 481 | 534 | 549 |
| 37 | 24 | 31 |  |  | 37 | 2880 | 1244 | 835 | 580 |
| 38 | 280 | 31 |  | 19 | 38 | 3226 | 963 | 1013 | 656 |
| 39 | 154 | 16 |  | 29 | 39 | 3180 | 1625 | 671 | 610 |
| 40 | 398 | 16 |  | 89 | 40 | 4102 | 1946 | 342 | 610 |
| 41 | 567 | 78 |  | 224 | 41 | 3572 | 2187 | 205 | 366 |
| 42 | 617 | 126 | 18 | 78 | 42 | 3318 | 1805 | 205 | 427 |
| 43 | 646 | 47 | 62 | 239 | 43 | 3433 | 2066 | 137 | 320 |
| 44 | 1192 | 180 | 204 | 245 | 44 | 2973 | 2367 | 246 | 244 |
| 45 | 847 | 184 | 248 | 270 | 45 | 2419 | 2227 | 178 | 122 |
| 46 | 1642 | 159 | 254 | 310 | 46 | 1728 | 2628 | 233 | 153 |
| 47 | 1345 | 264 | 376 | 144 | 47 | 2327 | 2628 | 288 | 76 |
| 48 | 1100 | 254 | 752 | 300 | 48 | 1959 | 2086 | 411 | 183 |
| 49 | 1669 | 279 | 924 | 94 | 49 | 2143 | 1344 | 425 | 198 |
| 50 | 1548 | 763 | 1114 | 176 | 50 | 1360 | 1705 | 260 | 137 |
| 51 | 1770 | 995 | 1648 | 399 | 51 | 346 | 622 | 0 | 15 |
| 52 | 1701 | 1138 | 1967 | 327 | 52 | 138 | 782 | 0 | 76 |
| 53 | 1623 | 822 | 2466 | 379 | 53 | 69 | 582 | 0 | 76 |
| 54 | 1224 | 1111 | 2394 | 668 | 54 | 69 | 461 | 0 | 15 |
| 55 | 935 | 1072 | 2116 | 459 | 55 | 115 | 582 | 0 | 107 |
| 56 | 1355 | 1103 | 2288 | 390 | 56 | 46 | 140 | 0 | 92 |


| 2008 | Retained |  |  |  | Discarded |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | FR-GADOID |  |  |  | FR-GADOID |  |  |  |  |
|  | 2008-Q1 | 2008 - Q2 | 2008 - Q3 | 2008-Q4 |  | 2008-Q1 | 2008 - Q2 | 2008-Q3 | 2008-Q4 |
| Length (cm) |  |  |  |  | Length (cm) |  |  |  |  |
| 57 | 824 | 1054 | 2415 | 702 | 57 | 46 | 181 | 0 | 107 |
| 58 | 907 | 832 | 2151 | 509 | 58 | 46 | 140 | 0 | 107 |
| 59 | 1087 | 995 | 2003 | 882 | 59 | 0 | 60 | 0 | 137 |
| 60 | 615 | 864 | 1999 | 871 | 60 | 0 | 120 | 0 | 61 |
| 61 | 1344 | 822 | 1768 | 555 | 61 | 23 | 0 | 0 | 92 |
| 62 | 948 | 731 | 1409 | 599 | 62 | 0 | 0 | 0 | 15 |
| 63 | 1372 | 765 | 1292 | 920 | 63 | 0 | 20 | 0 | 61 |
| 64 | 1283 | 669 | 1236 | 815 | 64 | 23 | 0 | 0 | 0 |
| 65 | 1410 | 633 | 1042 | 1209 | 65 |  |  |  | 31 |
| 66 | 1506 | 750 | 1480 | 639 | 66 |  |  |  | 0 |
| 67 | 2435 | 890 | 703 | 744 | 67 |  |  |  | 0 |
| 68 | 2504 | 1014 | 880 | 774 | 68 |  |  |  | 15 |
| 69 | 2038 | 796 | 623 | 691 | 69 |  |  |  | 0 |
| 70 | 2142 | 1067 | 542 | 474 | 70 |  |  |  | 31 |
| 71 | 2055 | 1020 | 532 | 445 | 71 |  |  |  | 46 |
| 72 | 2024 | 978 | 625 | 304 | 72 |  |  |  | 61 |
| 73 | 1885 | 1065 | 638 | 271 | 73 |  |  |  | 15 |
| 74 | 1664 | 869 | 705 | 135 | 74 |  |  |  | 31 |
| 75 | 1398 | 963 | 609 | 128 | 75 |  |  |  | 15 |

Table 7.2.4c. Continued.


| 2008 R | Retained |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | FR-GADOID |  |  |  | FR-GADOID |  |  |  |
|  | 2008-Q1 2008-Q2 2008-Q3 2008-Q4 |  |  |  | 008- Q1 | 2008-Q2 | 2008-Q3 | 2008-Q4 |
| Length <br> (cm) |  |  |  |  |  |  |  |  |
| 115 |  |  |  |  |  |  |  |  |
| Tot N | 61411 | 33681 | 50084 | 19810 | 46639 | 32438 | 8846 | 8632 |
| N trips sampled | 5 | 13 | 10 | 6 |  |  |  |  |
| N hauls sampled | 91 | 288 | 154 | 67 |  |  |  |  |

Table 7.2.4c. Cod in Divisions VIIe-k. Length structure of French landings and discards from the self-sampling Program Sampling data of FR-GADOID raised by landing ratio to the total catch of the fleet in VIIfgh assuming the same discarding practice as the French Gadoid trawlers.


| 2008 | Retained |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | FR-NEPHROPS |  |  |  | FR-NEPHR | ROPS |  |  |
|  | 2008-Q1 2008-Q2 | 2008-Q3 | 2008 - Q4 |  | 2008-Q1 | 2008 - Q2 | 2008-Q3 | 2008-Q4 |
| Length (cm) |  |  |  | Length (cm) |  |  |  |  |
| 46 | 1760 | 92 | 133 | 46 | 497 | 5115 | 190 | 189 |
| 47 | 10686 | 154 | 218 | 47 | 670 | 5115 | 235 | 94 |
| 48 | 196191 | 395 | 143 | 48 | 563 | 4061 | 335 | 226 |
| 49 | $288 \quad 211$ | 167 | 79 | 49 | 616 | 2616 | 346 | 245 |
| 50 | $144 \quad 359$ | 0 | 55 | 50 | 391 | 3319 | 212 | 170 |
| 51 | 255485 | 228 | 86 | 51 | 99 | 1211 | 0 | 19 |
| 52 | 163932 | 1052 | 103 | 52 | 40 | 1523 | 0 | 94 |
| 53 | 206651 | 281 | 147 | 53 | 20 | 1132 | 0 | 94 |
| 54 | 3821378 | 977 | 176 | 54 | 20 | 898 | 0 | 19 |
| 55 | 1251591 | 754 | 113 | 55 | 33 | 1132 | 0 | 132 |
| 56 | 351381 | 1027 | 227 | 56 | 13 | 273 | 0 | 113 |
| 57 | 1531756 | 999 | 171 | 57 | 13 | 351 | 0 | 132 |
| 58 | 230864 | 797 | 384 | 58 | 13 | 273 | 0 | 132 |
| 59 | $145 \quad 2882$ | 1229 | 183 | 59 | 0 | 117 | 0 | 170 |
| 60 | $91 \quad 2178$ | 1169 | 251 | 60 | 0 | 234 | 0 | 75 |
| 61 | 1321092 | 701 | 449 | 61 | 7 | 0 | 0 | 113 |
| 62 | 129584 | 1606 | 585 | 62 | 0 | 0 | 0 | 19 |
| 63 | $76 \quad 292$ | 1116 | 627 | 63 | 0 | 39 | 0 | 75 |
| 64 | 263 2570 | 633 | 565 | 64 | 7 | 0 | 0 | 0 |
| 65 | 163106 | 745 | 318 | 65 | 0 | 0 | 0 | 38 |
| 66 | 3231083 | 252 | 113 | 66 | 0 | 0 | 0 | 0 |
| 67 | 250791 | 648 | 111 | 67 | 0 | 0 | 0 | 0 |
| 68 | 4210 | 952 | 654 | 68 | 0 | 0 | 0 | 19 |
| 69 | $249 \quad 487$ | 706 | 210 | 69 | 0 | 0 | 0 | 0 |
| 70 | $340 \quad 2931$ | 764 | 205 | 70 | 0 | 0 | 0 | 38 |

Table 7.2.4c. Continued.



Table 7.2.4d. Cod in Divisions VIIe-k. Length structure of French landings and discards from onboard observer program. Otter Trawlers targeting demersal fish.

|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_DEF |  |  |  |  | OT_DEF |  |  |  |
| Length (cm) | $\begin{aligned} & 2009 \text { - } \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length (cm) | $2009 \text { - Q1 }$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ |
| 20 | 0 | 0 | 0 | 0 | 20 | 0 | 0 | 0 | 0 |
| 21 | 0 | 0 | 0 | 0 | 21 | 0 | 0 | 0 | 0 |
| 22 | 0 | 0 | 0 | 0 | 22 | 0 | 0 | 0 | 0 |
| 23 | 0 | 0 | 0 | 0 | 23 | 0 | 0 | 0 | 0 |
| 24 | 0 | 0 | 0 | 0 | 24 | 0 | 0 | 0 | 0 |
| 25 | 0 | 0 | 0 | 0 | 25 | 0 | 0 | 0 | 0 |
| 26 | 0 | 0 | 0 | 0 | 26 | 0 | 0 | 0 | 0 |
| 27 | 0 | 0 | 0 | 0 | 27 | 0 | 0 | 0 | 4 |
| 28 | 0 | 0 | 0 | 0 | 28 | 0 | 0 | 0 | 0 |
| 29 | 0 | 0 | 0 | 0 | 29 | 0 | 0 | 1 | 0 |
| 30 | 0 | 0 | 0 | 0 | 30 | 0 | 0 | 0 | 0 |
| 31 | 0 | 0 | 0 | 0 | 31 | 0 | 0 | 3 | 8 |
| 32 | 0 | 0 | 0 | 0 | 32 | 0 | 0 | 7 | 4 |
| 33 | 0 | 0 | 0 | 0 | 33 | 0 | 0 | 2 | 21 |
| 34 | 0 | 0 | 0 | 0 | 34 | 0 | 0 | 42 | 28 |
| 35 | 0 | 0 | 1 | 0 | 35 | 0 | 0 | 22 | 17 |
| 36 | 0 | 0 | 3 | 0 | 36 | 0 | 0 | 36 | 32 |
| 37 | 0 | 0 | 7 | 0 | 37 | 0 | 0 | 28 | 25 |
| 38 | 0 | 0 | 9 | 0 | 38 | 0 | 0 | 32 | 36 |
| 39 | 0 | 0 | 12 | 0 | 39 | 0 | 0 | 27 | 26 |
| 40 | 0 | 0 | 7 | 2 | 40 | 0 | 0 | 31 | 22 |
| 41 | 0 | 0 | 8 | 0 | 41 | 0 | 0 | 22 | 37 |
| 42 | 0 | 0 | 6 | 12 | 42 | 0 | 0 | 24 | 30 |
| 43 | 0 | 0 | 3 | 10 | 43 | 0 | 0 | 12 | 16 |
| 44 | 0 | 0 | 5 | 4 | 44 | 0 | 0 | 15 | 8 |
| 45 | 0 | 0 | 5 | 8 | 45 | 0 | 0 | 5 | 4 |
| 46 | 0 | 0 | 4 | 6 | 46 | 0 | 0 | 3 | 0 |
| 47 | 0 | 0 | 3 | 16 | 47 | 0 | 0 | 0 | 0 |
| 48 | 0 | 0 | 2 | 7 | 48 | 0 | 0 | 2 | 0 |
| 49 | 0 | 0 | 5 | 6 | 49 | 0 | 0 | 0 | 0 |
| 50 | 0 | 0 | 1 | 5 | 50 | 0 | 0 | 1 | 0 |
| 51 | 0 | 0 | 3 | 2 | 51 | 0 | 0 | 1 | 0 |
| 52 | 0 | 1 | 2 | 4 | 52 | 0 | 0 | 0 | 0 |
| 53 | 0 | 1 | 8 | 2 | 53 | 0 | 0 | 0 | 0 |
| 54 | 0 | 0 | 16 | 2 | 54 | 0 | 0 | 0 | 0 |
| 55 | 0 | 0 | 17 | 2 | 55 | 0 | 0 | 0 | 0 |
| 56 | 0 | 0 | 8 | 4 | 56 | 0 | 0 | 0 | 0 |
| 57 | 0 | 0 | 6 | 5 | 57 | 0 | 0 | 0 | 0 |
| 58 | 0 | 0 | 19 | 6 | 58 | 0 | 0 | 0 | 0 |


|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_DEF |  |  |  |  | OT_DEF |  |  |  |
| Length (cm) | $\begin{aligned} & 2009- \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length (cm) | 2009- Q1 | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009 \\ & \text { Q4 } \end{aligned}$ |
| 59 | 0 | 1 | 8 | 8 | 59 | 0 | 0 | 0 | 0 |
| 60 | 0 | 0 | 15 | 5 | 60 | 0 | 0 | 0 | 0 |
| 61 | 0 | 1 | 19 | 1 | 61 | 0 | 0 | 0 | 0 |
| 62 | 0 | 1 | 37 | 10 | 62 | 0 | 0 | 0 | 0 |
| $63$ | $0$ | 0 | $18$ | 7 | 63 | 0 | 0 | 0 | 0 |
| 64 | 0 | 0 | 31 | 6 | 64 | 0 | 0 | 0 | 0 |
| 65 | 0 | 0 | 13 | 7 | 65 | 0 | 0 | 0 | 0 |
| $66$ | $0$ | $0$ | $14$ | $13$ | 66 | 0 | 0 | 0 | 0 |
| 67 | $0$ | $1$ | $11$ | 5 | 67 | $0$ | 0 | 0 | 0 |
| 68 | 0 | 0 | 10 | 7 | 68 | 0 | 0 | 0 | 0 |
| 69 | 0 | 0 | 11 | 5 | 69 | 0 | 0 | 0 | 0 |
| 70 | 0 | 0 | 10 | 6 | 70 | 0 | 0 | 0 | 0 |

Table 7.2.4d. Continued.

|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_DEF |  |  |  |  | OT_DEF |  |  |  |
| Length (cm) | $\begin{aligned} & 2009- \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length (cm) | $2009 \text { - Q1 }$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ |
| 71 | 0 | 1 | 11 | 5 | 71 | 0 | 0 | 0 | 0 |
| 72 | 0 | 2 | 4 | 2 | 72 | 0 | 0 | 0 | 0 |
| 73 | 0 | 0 | 4 | 1 | 73 | 0 | 0 | 0 | 0 |
| 74 | 0 | 2 | 7 | 1 | 74 | 0 | 0 | 0 | 0 |
| 75 | 0 | 2 | 5 | 3 | 75 | 0 | 0 | 0 | 0 |
| 76 | 0 | 0 | 8 | 0 | 76 | 0 | 0 | 0 | 0 |
| 77 | 0 | 1 | 7 | 2 | 77 | 0 | 0 | 0 | 0 |
| 78 | 0 | 0 | 4 | 2 | 78 | 0 | 0 | 0 | 0 |
| 79 | 0 | 1 | 9 | 2 | 79 | 0 | 0 | 0 | 0 |
| 80 | 0 | 3 | 13 | 0 | 80 | 0 | 0 | 0 | 0 |
| 81 | 0 | 0 | 8 | 1 | 81 | 0 | 0 | 0 | 0 |
| 82 | 0 | 2 | 8 | 1 | 82 | 0 | 0 | 0 | 0 |
| 83 | 0 | 1 | 14 | 1 | 83 | 0 | 0 | 0 | 0 |
| 84 | 0 | 0 | 5 | 1 | 84 | 0 | 0 | 0 | 0 |
| 85 | 0 | 1 | 9 | 0 | 85 | 0 | 0 | 0 | 0 |
| 86 | 0 | 0 | 7 | 2 | 86 | 0 | 0 | 0 | 0 |
| 87 | 0 | 2 | 3 | 0 | 87 | 0 | 0 | 0 | 0 |
| 88 | 0 | 2 | 2 | 0 | 88 | 0 | 0 | 0 | 0 |
| 89 | 0 | 1 | 6 | 0 | 89 | 0 | 0 | 0 | 0 |
| 90 | 0 | 0 | 4 | 0 | 90 | 0 | 0 | 0 | 0 |
| 91 | 0 | 1 | 1 | 0 | 91 | 0 | 0 | 0 | 0 |
| 92 | 0 | 0 | 2 | 0 | 92 | 0 | 0 | 0 | 0 |
| 93 | 0 | 1 | 2 | 0 | 93 | 0 | 0 | 0 | 0 |
| 94 | 0 | 0 | 0 | 0 | 94 | 0 | 0 | 0 | 0 |



Table 7.2.4d. Cod in Divisions VIIe-k. Length structure of French landings and discards from onboard observer programme. Otter Trawlers targeting Nephrops.

|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_CRU |  |  |  |  | OT_CRU |  |  |  |
| Length (cm) | $\begin{aligned} & 2009- \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length (cm) | 2009 - Q1 | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ |
| 20 | 0 | 0 | 0 | 0 | 20 | 0 | 0 | 0 | 0 |
| 21 | 0 | 0 | 0 | 0 | 21 | 0 | 0 | 0 | 0 |
| 22 | 0 | 0 | 0 | 0 | 22 | 0 | 0 | 0 | 0 |
| 23 | 0 | 0 | 0 | 0 | 23 | 0 | 0 | 0 | 0 |
| 24 | 0 | 0 | 0 | 0 | 24 | 0 | 0 | 0 | 0 |
| 25 | 0 | 0 | 0 | 0 | 25 | 0 | 0 | 0 | 0 |
| 26 | 0 | 0 | 0 | 0 | 26 | 0 | 0 | 0 | 0 |
| 27 | 0 | 0 | 0 | 0 | 27 | 0 | 27 | 0 | 0 |
| 28 | 0 | 0 | 0 | 0 | 28 | 0 | 8 | 18 | 0 |
| 29 | 0 | 0 | 0 | 0 | 29 | 0 | 0 | 1 | 0 |
| 30 | 0 | 0 | 0 | 0 | 30 | 0 | 4 | 0 | 0 |
| 31 | 0 | 0 | 3 | 0 | 31 | 0 | 0 | 12 | 0 |
| 32 | 0 | 0 | 0 | 0 | 32 | 0 | 42 | 157 | 0 |
| 33 | 0 | 0 | 3 | 0 | 33 | 0 | 12 | 0 | 0 |
| 34 | 0 | 0 | 3 | 0 | 34 | 0 | 0 | 107 | 0 |
| 35 | 0 | 1 | 44 | 0 | 35 | 0 | 3 | 14 | 0 |
| 36 | 0 | 0 | 75 | 0 | 36 | 0 | 43 | 34 | 0 |
| 37 | 0 | 0 | 16 | 0 | 37 | 0 | 7 | 28 | 0 |
| 38 | 0 | 1 | 48 | 0 | 38 | 0 | 3 | 18 | 0 |
| 39 | 0 | 2 | 11 | 0 | 39 | 0 | 0 | 19 | 0 |
| 40 | 0 | 0 | 5 | 0 | 40 | 0 | 0 | 0 | 0 |
| 41 | 0 | 1 | 12 | 0 | 41 | 0 | 0 | 0 | 0 |
| 42 | 0 | 1 | 4 | 0 | 42 | 0 | 0 | 16 | 0 |
| 43 | 0 | 1 | 3 | 0 | 43 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 2 | 0 | 44 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 4 | 0 | 45 | 0 | 0 | 0 | 0 |
| 46 | 0 | 0 | 3 | 0 | 46 | 0 | 0 | 0 | 0 |
| 47 | 0 | 0 | 5 | 0 | 47 | 0 | 0 | 0 | 0 |
| 48 | 0 | 1 | 0 | 0 | 48 | 0 | 0 | 0 | 0 |
| 49 | 0 | 3 | 0 | 0 | 49 | 0 | 0 | 0 | 0 |
| 50 | 0 | 2 | 0 | 0 | 50 | 0 | 0 | 0 | 0 |
| 51 | 0 | 4 | 1 | 0 | 51 | 0 | 0 | 0 | 0 |
| 52 | 0 | 5 | 3 | 0 | 52 | 0 | 0 | 0 | 0 |
| 53 | 0 | 4 | 3 | 0 | 53 | 0 | 0 | 0 | 0 |
| 54 | 0 | 7 | 4 | 0 | 54 | 0 | 0 | 0 | 0 |
| 55 | 0 | 10 | 4 | 0 | 55 | 0 | 0 | 0 | 0 |
| 56 | 0 | 7 | 5 | 0 | 56 | 0 | 0 | 0 | 0 |
| 57 | 0 | 6 | 7 | 0 | 57 | 0 | 0 | 0 | 0 |
| 58 | 0 | 10 | 10 | 0 | 58 | 0 | 0 | 0 | 0 |


|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_CRU |  |  |  |  | OT_CRU |  |  |  |
| Length (cm) | $\begin{aligned} & 2009- \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length (cm) | 2009 - Q1 | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009 \\ & \text { Q4 } \end{aligned}$ |
| 59 | 0 | 7 | 6 | 0 | 59 | 0 | 0 | 0 | 0 |
| 60 | 0 | 6 | 10 | 0 | 60 | 0 | 0 | 0 | 0 |
| 61 | 0 | 12 | 3 | 0 | 61 | 0 | 0 | 0 | 0 |
| 62 | 0 | 6 | 8 | 0 | 62 | 0 | 0 | 0 | 0 |
| 63 | 0 | 4 | 6 | 0 | 63 | 0 | 0 | 0 | 0 |
| 64 | 0 | 9 | 5 | 0 | 64 | 0 | 0 | 0 | 0 |
| 65 | 0 | 9 | 6 | 0 | 65 | 0 | 0 | 0 | 0 |
| 66 | 0 | $12$ | 7 | 0 | 66 | 0 | 0 | 0 | 0 |
| 67 | 0 | $7$ | 7 | 0 | 67 | 0 | 0 | 0 | 0 |
| 68 | 0 | 16 | 3 | 0 | 68 | 0 | 0 | 0 | 0 |
| 69 | 0 | 12 | 2 | 0 | 69 | 0 | 0 | 0 | 0 |
| 70 | 0 | 18 | 9 | 0 | 70 | 0 | 0 | 0 | 0 |

Table 7.2.4d. Continued.

|  | Retained |  |  |  |  | Discarded |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | OT_CRU |  |  |  |  | OT_CRU |  |  |  |
| Length (cm) | $\begin{aligned} & 2009- \\ & \text { Q1 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ | Length <br> (cm) | $2009 \text { - Q1 }$ | $\begin{aligned} & 2009- \\ & \text { Q2 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q3 } \end{aligned}$ | $\begin{aligned} & 2009- \\ & \text { Q4 } \end{aligned}$ |
| 71 | 0 | 18 | 9 | 0 | 71 | 0 | 0 | 0 | 0 |
| 72 | 0 | 30 | 3 | 0 | 72 | 0 | 0 | 0 | 0 |
| 73 | 0 | 17 | 6 | 0 | 73 | 0 | 0 | 0 | 0 |
| 74 | 0 | 19 | 3 | 0 | 74 | 0 | 0 | 0 | 0 |
| 75 | 0 | 27 | 4 | 0 | 75 | 0 | 0 | 0 | 0 |
| 76 | 0 | 19 | 8 | 0 | 76 | 0 | 0 | 0 | 0 |
| 77 | 0 | 35 | 6 | 0 | 77 | 0 | 0 | 0 | 0 |
| 78 | 0 | 20 | 7 | 0 | 78 | 0 | 0 | 0 | 0 |
| 79 | 0 | 16 | 6 | 0 | 79 | 0 | 0 | 0 | 0 |
| 80 | 0 | 21 | 6 | 0 | 80 | 0 | 0 | 0 | 0 |
| 81 | 0 | 11 | 3 | 0 | 81 | 0 | 0 | 0 | 0 |
| 82 | 0 | 18 | 5 | 0 | 82 | 0 | 0 | 0 | 0 |
| 83 | 0 | 18 | 8 | 0 | 83 | 0 | 0 | 0 | 0 |
| 84 | 0 | 13 | 4 | 0 | 84 | 0 | 0 | 0 | 0 |
| 85 | 0 | 16 | 3 | 0 | 85 | 0 | 0 | 0 | 0 |
| 86 | 0 | 11 | 3 | 0 | 86 | 0 | 0 | 0 | 0 |
| 87 | 0 | 14 | 1 | 0 | 87 | 0 | 0 | 0 | 0 |
| 88 | 0 | 9 | 1 | 0 | 88 | 0 | 0 | 0 | 0 |
| 89 | 0 | 8 | 2 | 0 | 89 | 0 | 0 | 0 | 0 |
| 90 | 0 | 19 | 3 | 0 | 90 | 0 | 0 | 0 | 0 |
| 91 | 0 | 15 | 4 | 0 | 91 | 0 | 0 | 0 | 0 |
| 92 | 0 | 3 | 2 | 0 | 92 | 0 | 0 | 0 | 0 |
| 93 | 0 | 10 | 2 | 0 | 93 | 0 | 0 | 0 | 0 |
| 94 | 0 | 8 | 5 | 0 | 94 | 0 | 0 | 0 | 0 |



Table 7.2.5. Cod in Divisions VIIe-k. Catch weight-at-age.

Run title : Cod in Divisions VIIe-k, WGCSE10, index file

At 6/05/2010 12:02

Table 2 Catch weights at age (kg)
YEAR, 1971, 1972, 1973, 1974, 1975, 1976, 1977, 1978, 1979,

AGE

| .9080, | .9080, | . 9080, | .9080, | .9080, | .9080, | . 9080, | 9080, | 9080, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2.1930, | 2.1930, | 2.1930, | 2.1930, | 2.1930, | 2.1930, | 2.1930, | 2.1930, | 2.1930, |
| 4.8310, | 4.8310, | 4.8310, | 4.8310, | 4.8310, | 4.8310, | 4.8310, | 4.8310, | 4.8310, |
| 7.4640, | 7.4640, | 7.4640, | 7.4640, | 7.4640, | 7.4640, | 7.4640, | 7.4640, | 7.4640, |
| 9.6690, | 9.6690, | 9.6690, | 9.669 | 9.6690, | 9.6690, | 9.6690, | 9.6690, | 9.6690, |
| 11.7840, | 11.7840, | 11.7840 | 1.7840 | 11.7840 | 11.7840 | 11.7840 | 11.7840 | 11.7840, |
| 14.8159, | 14.4792, | . 6675 | .9506, | 4.5262, | .127 | . 714 | .2267, | 14.3395, |
| 1.0006, | .9972, | .9982, | .9966, | 1.0011, | 1.0029, | 1.0004, | .9974, | 1.0006, |

Table 2 Catch weights at age (kg)
YEAR, 1980, 1981, 1982, 1983, 1984, 1985, 1986, 1987, 1988, 1989,

AGE
1, .9080, .9450, .9450, .9790, .9810, 1.0010, 1.0540, .9090, .9060, .8440,
2, 2.1930, 1.5490, 2.2420, 2.5250, 2.6450, 2.6370, 2.5540, 2.5040, 2.1870, 2.0130,
3, 4.8310, 4.3850, 4.4740, 4.9610, 5.2840, 5.5210, 5.3980, 5.2640, 5.3180, 4.7060,
4, 7.4640, 7.5650, 7.7970, 7.4570, 7.8280, 8.0820, 7.4400, 8.0890, 7.9970, 7.6380,
5, $9.6690,9.0600,10.2500,9.9650,9.7580,10.4070,10.7820,10.4470,10.6490,9.4380$,
6, $11.7840,12.7500,12.4650,12.0100,11.6720,11.4690,12.3960,13.5740,12.4860,12.9170$,
+gp, $13.8620,14.7237,15.4408,16.4710,15.3396,14.3697,13.5580,15.3490,14.6217,13.3935$,
0 SOPCOFAC, $1.0003,1.0002,1.0146,1.0006, .9984,1.0092,1.0000, .9844, .9997,1.0003$,

Run title : Cod in Divisions VIIe-k,WGCSE10,index file

```
At 6/05/2010 12:02
```

Table 2 Catch weights at age (kg)
YEAR, 1990, 1991, 1992, 1993, 1994, 1995, 1996, 1997, 1998, 1999,

AGE
1, .8800, .9050, .8150, .8710, .8740, .8060, .7870, .7710, .8530, .9930,
2, 2.3000, 2.1350, 1.9160, 2.0430, 2.0000, 1.9730, 1.8770, 2.0390, 1.8960, 2.0980,
3, 4.6240, 4.9870, 4.9160, 4.5080, 4.4920, 4.5890, 4.6390, 4.5160, 4.4610, 4.4950,
4, 7.1880, 6.7380, 7.3590, 6.8660, 7.9260, 7.5600, 6.9970, 7.3890, 6.8810, 7.3260,
5, 9.0450, 8.8650, 9.7440, 8.4310, 10.0920, 9.7500, 9.8540, 9.7190, 9.3290, 8.9450,
6, 11.7130, 10.8090, 11.4980, 10.9420, 12.2120, 11.1520, 11.4070, 11.8200, 11.2160, 11.2550,
+gp, $14.8144,14.1344,12.6295,12.3344,14.0578,14.0814,12.3707,14.3670,14.0713,14.6309$,
SOPCOFAC, .9900, 1.0000, 1.0000, 1.0009, 1.0000, .9999, 1.0000, 1.0006, 1.0012, 1.0017,


Table 7.2.6. Cod in Divisions VIIe-k. Stock weight-at-ages = 1st quarter values.

Run title : Cod in Divisions VIIe-k,WGCSE10, index file

At 6/05/2010 12:02

| Table | Stock | ights at | age (kg) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1971, | 1972, | 1973, | 1974, | 1975, | 1976, | 1977, | 1978, | 1979, |
| AGE |  |  |  |  |  |  |  |  |  |
| 1, | .6620, | .6620, | .6620, | .6620, | .6620, | .6620, | .6620, | .6620, | .6620, |
| 2, | 1.7090, | 1.7090, | 1.7090, | 1.7090, | 1.7090, | 1.7090, | 1.7090, | 1.7090, | 1.7090, |
| 3, | 4.4440, | 4.4440, | 4.4440, | 4.4440, | 4.4440, | 4.4440, | 4.4440, | 4.4440, | 4.4440, |
| 4, | 7.3210, | 7.3210, | 7.3210, | 7.3210, | 7.3210, | 7.3210, | 7.3210, | 7.3210, | 7.3210, |
| 5, | 9.5290, | 9.5290, | 9.5290, | 9.5290, | 9.5290, | 9.5290, | 9.5290, | 9.5290, | 9.5290, |
| 6, | 11.6050, | 11.6050, | 11.6050, | 11.6050, | 11.6050, | 11.6050, | 11.6050, | 11.6050, | 11.6050, |
| +gp, | 14.5404, | 14.1778, | 14.3755, | 14.5822, | 14.2402, | 14.8683, | 15.3589, | 14.9079, | 14.0056, |

Table 3 Stock weights at age (kg)
YEAR, 1980, 1981, 1982, 1983, 1984, 1985, 1986, 1987, 1988, 1989,

AGE
1, .6620, .4600, .7040, .4460, .5120, .5810, .5280, .5220, .9060, .8440,
2, 1.7090, 1.5490, 1.4880, 1.9450, 1.9510, 2.0700, 1.9020, 1.9470, 1.6210, 1.4630,
3, 4.4440, 2.2840, 3.8760, 4.4670, 4.9280, 5.3330, 5.2860, 4.8770, 4.8870, 4.5140,
4, 7.3210, 7.8060, 7.4070, 7.3530, 7.4330, 8.3760, 7.3820, 7.9460, 7.7770, 7.6150,
5, 9.5290, 10.5440, 9.6240, 9.7520, 9.5520, 10.8510, 10.6890, 10.3080, 10.3020, 9.4380,
6, 11.6050, 11.4390, 12.3160, 11.2230, 12.1800, 11.5850, 12.3930, 14.4190, 11.7860, 12.6920, +gp, $13.5130,14.6123,15.7394,17.4511,15.2018,14.9743,14.4820,15.4457,13.4600,14.1533$,

Table 3 Stock weights at age (kg)
YEAR, 1990, 1991, 1992, 1993, 1994, 1995, 1996, 1997, 1998, 1999,

AGE
1, .6130, .5390, .6630, .7030, .6050, .6120, .6730, .4700, .4210, .7780,
2, $1.7740,1.5380,1.3180,1.3850,1.7540,1.4440,1.2830,1.4100,1.3140,1.5420$,
3, 4.3900, 4.7910, 4.6000, 4.2780, 4.1890, 4.3460, 4.4710, 4.0790, 4.3400, 4.2520,
4, 7.1860, 6.5240, 6.5580, 6.5740, 7.7200, 7.4520, 6.7470, 7.1120, 6.6760, 7.1260,
5, 8.4860, 8.6310, 9.3420, 8.0660, 9.7220, 9.1400, 9.8770, 9.0440, 9.3030, 8.7000,
$6,10.7030,10.6720,11.2850,10.8150,12.1010,10.6460,11.4240,11.1560,11.1720,11.1420$,
+gp, $14.6578,13.8090,12.4660,12.1295,13.9081,14.0514,12.8480,13.7300,12.8280,15.2226$,
Table 3 Stock weights at age (kg)
YEAR,
AGE

Table 7.2.7. Cod in Divisions VIIe-k. Series of surveys indices scrutinized at WGCSE.

UK-WCGFS West Coast March survey, effort in mn towed, numbers *10**2, final survey in 2004

| 1992 | 2004 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 1 | 0.15 | 0.25 |  |  |
| 1 | 5 |  |  |  |  |
| 3774 | 2800 | 7100 | 400 | 200 | 200 |
| 3602 | 500 | 7250 | 4850 | 1230 | 100 |
| 1915 | 7400 | 600 | 3180 | 1130 | 300 |
| 3439 | 11200 | 14520 | 880 | 1400 | 700 |
| 3695 | 1300 | 6800 | 8500 | 1000 | 800 |
| 3826 | 3700 | 3200 | 3400 | 700 | 100 |
| 3744 | 1800 | 2500 | 2000 | 700 | 500 |
| 3823 | 200 | 1500 | 300 | 400 | 100 |
| 4092 | 3000 | 0 | 410 | 200 | 200 |
| 3700 | 1450 | 1100 | 1000 | 100 | 100 |
| 3387 | 200 | 5450 | 2960 | 430 | 100 |
| 2326 | 0 | 579 | 3154 | 410 | 100 |
| 1689 | 1400 | 0 | 200 | 1000 | 200 |

FR-EVHOE Groundfish Oct-Nov survey in VIIf, $\mathbf{g}, \mathrm{h}, \mathbf{j}$, numbers per 30 mn 19972009

| 1 | 1 | 0.75 | 1 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 6 |  |  |  |  |  |  |
| 1 | 0.213 | 0.095 | 0.246 | 0.117 | 0.048 | 0 | 1997 |
| 1 | 0.212 | 0.52 | 0.207 | 0.045 | 0.045 | 0 | 1998 |
| 1 | 0.155 | 0.184 | 0.283 | 0.015 | 0.03 | 0.015 | 1999 |
| 1 | 1.046 | 0.041 | 0.118 | 0.064 | 0.013 | 0 | 2000 |
| 1 | 0.716 | 0.18 | 0.029 | 0.038 | 0.018 | 0.007 | 2001 |
| 1 | 0.033 | 0.313 | 0.148 | 0 | 0.015 | 0 | 2002 |
| 1 | 0.052 | 0.041 | 0.142 | 0.061 | 0.008 | 0 | 2003 |
| 1 | 0.066 | 0.144 | 0.072 | 0.122 | 0.046 | 0 | 2004 |
| 1 | 0.255 | 0.12 | 0.055 | 0 | 0.026 | 0 | 2005 |
| 1 | 0.125 | 0.139 | 0 | 0.048 | 0.045 | 0 | 2006 |
| 1 | 0.321 | 0.206 | 0.117 | 0.033 | 0 | 0 | 2007 |
| 1 | 0.217 | 0.141 | 0.117 | 0.096 | 0 | 0 | 2008 |
| 1 | 0.237 | 0.092 | 0.132 | 0.078 | 0 | 0.023 | 2009 |

IR-GFS-VIIgj combined: Irish Grounfish Survey (IBTS 4th Qrt)- Cod number per 30 mn towed (Interim indices for the new Celtic Explorer series)
20032009

| 1 | 1 | 0.79 | 0.92 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 6 |  |  |  | 0 | 2003 |  |
| 1 | 0.167 | 0.223 | 0.229 | 0.075 | 0 | 0 | 2004 |
| 1 | 0.3 | 0.106 | 0.035 | 0.018 | 0.018 | 0 | 2005 |
| 1 | 0.967 | 0.138 | 0.035 | 0 | 0 | 0 | 2006 |
| 1 | 0.632 | 0.2 | 0.031 | 0 | 0 | 0.015 | 2007 |
| 1 | 0.837 | 0.279 | 0.103 | 0.029 | 0 | 0 | 2008 |
| 1 | 0.164 | 0.432 | 0.104 | 0.015 | 0 | 0 | 2009 |

Table 7.2.8a. Cod in Divisions VIIe-k. Time-series of landings, effort and lpue.

|  | France |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Fr gadoid trawlers VIIfgh |  |  | Fr Nephrops trawlers VIIfgh |  |  | Fr Otter trawlers VIIe-k |  |  | Fr Otter trawlers VII e |  |  |
| Year | Landings | Effor |  | Landings Effort lpue |  |  | Landings Effort |  | lpue | Landings | Effort lpue |  |
| 1978 | Q2+Q3+Q4 for |  |  | Q2+Q3+Q4 for |  |  |  |  |  |  |  |  |
| 1979 | consistenc with |  |  | consisten <br> with |  |  | includes Fr trawlers | Fr gadoid and |  |  |  |  |
| 1980 | box closure |  |  | box closure |  |  | Fr Nephrop trawlers |  |  |  |  |  |
| 1981 | during | 2005 |  | during | 2005 |  |  |  |  |  |  |  |
| 1982 | $\begin{aligned} & \text { and Feb-I } \\ & 2008 \end{aligned}$ | arch 2 | 06 to | and Feb- | March | $06 \text { to }$ | $2008$ |  |  |  |  |  |
| 1983 | 1453 | 75.0 | 19.4 | 630 | 190.5 | 3.3 | 5443 | 904.3 | 6.0 | 472 | 210.6 | 2.2 |
| 1984 | 2002 | 60.6 | 33.1 | 671 | 170.5 | 3.9 | 4881 | 654.9 | 7.5 | 189 | 118.4 | 1.6 |
| 1985 | 1667 | 73.4 | 22.7 | 1023 | 150.7 | 6.8 | 6262 | 847.6 | 7.4 | 351 | 154.1 | 2.3 |
| 1986 | 2086 | 85.3 | 24.5 | 774 | 132.6 | 5.8 | 8046 | 932.0 | 8.6 | 431 | 220.4 | 2.0 |
| 1987 | 2804 | 107.8 | 26.0 | 778 | 145.7 | 5.3 | 8215 | 886.0 | 9.3 | 835 | 167.6 | 5.0 |
| 1988 | 6243 | 184.4 | 33.9 | 1726 | 144.1 | 12.0 | 13739 | 963.6 | 14.3 | 1320 | 199.4 | 6.6 |
| 1989 | 5171 | 166.3 | 31.1 | 1496 | 157.7 | 9.5 | 15715 | 1066.0 | 14.7 | 983 | 217.4 | 4.5 |
| 1990 | 3045 | 155.2 | 19.6 | 1138 | 206.3 | 5.5 | 9018 | 1073.3 | 8.4 | 383 | 198.6 | 1.9 |
| 1991 | 2096 | 127.1 | 16.5 | 690 | 186.2 | 3.7 | 5878 | 1013.2 | 5.8 | 335 | 177.7 | 1.9 |
| 1992 | 2304 | 133.0 | 17.3 | 1223 | 226.2 | 5.4 | 6709 | 1060.6 | 6.3 | 325 | 179.1 | 1.8 |
| 1993 | 2566 | 155.5 | 16.5 | 1236 | 205.3 | 6.0 | 8302 | 1095.6 | 7.6 | 295 | 238.4 | 1.2 |
| 1994 | 1725 | 121.8 | 14.2 | 1245 | 225.1 | 5.5 | 7353 | 959.7 | 7.7 | 306 | 185.1 | 1.7 |
| 1995 | 2598 | 128.2 | 20.3 | 1606 | 200.5 | 8.0 | 8248 | 1010.8 | 8.2 | 520 | 215.2 | 2.4 |
| 1996 | 2455 | 123.0 | 20.0 | 1450 | 181.6 | 8.0 | 8667 | 954.6 | 9.1 | 460 | 188.5 | 2.4 |
| 1997 | 2830 | 168.2 | 16.8 | 1246 | 152.6 | 8.2 | 8307 | 1057.5 | 7.9 | 584 | 258.3 | 2.3 |
| 1998 | 1707 | 139.3 | 12.3 | 805 | 111.1 | 7.2 | 5765 | 743.383* | 7.76* | 150* | 28.2* | 5.33* |
| 1999 | 1271 | 138.8 | 9.2 | 546 | 114.6 | 4.8 | 5445 | 1047.3 | 5.2 | 647 | 298.4 | 2.2 |
| 2000 | 938 | 115.3 | 8.1 | 711 | 125.3 | 5.7 | 4254 | 1051.9 | 4.0 | 542 | 312.5 | 1.7 |
| 2001 | 1911 | 138.5 | 13.8 | 916 | 141.7 | 6.5 | 5957 | 1010.4 | 5.9 | 584 | 281.3 | 2.1 |
| 2002 | 2412 | 121.8 | 19.8 | 1083 | 147.6 | 7.3 | 7389 | 974.8 | 7.6 | 654 | 317.4 | 2.1 |
| 2003 | 1110 | 92.0 | 12.1 | 972 | 169.9 | 5.7 | 5157 | 1025.7 | 5.0 | 619 | 366.2 | 1.7 |
| 2004 | 469 | 83.1 | 5.6 | 462 | 128.2 | 3.6 | 2379 | 952.1 | 2.4 | 193 | 353.6 | 0.5 |
| 2005 | 483 | 79.1 | 6.1 | 343 | 113.3 | 3.0 | 1577 | 874.2 | 1.7 | 239 | 333.9 | 0.7 |
| 2006 | 430 | 55.6 | 7.7 | 376 | 108.3 | 3.5 | 1834 | 866.8 | 2.1 | 359 | 334.8 | 1.1 |
| 2007 | 678 | 63.4 | 10.7 | 509 | 85.1 | 6.0 | 2438 | 805.7 | 3.0 | 445 | 311.5 | 1.4 |
| 2008 | 496 | 54.0 | 9.2 | 445 | 78.1 | 5.7 | 1958 | 655.3 | 3.0 | 399 | 242.5 | 1.6 |
| 2009 | data unav | ailable |  |  |  |  |  |  |  |  |  |  |

Units: landings in Tonnes live weight, Effort in 000s hours fished, lpue in $\mathrm{Kg} /$ hour fished

* unreliable


## Fr Nephrops trawlers

|  | Fr gadoid trawlers VIIfgh |  |  |  |  |  |  |  |  |  | VIIfgh |  | Fr Otter trawlers VIle-k |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings | Effort | lpue | Landings | Effort | lpue | Landings | Effort | lpue |  |  |  |  |  |
| FR- High-grading input |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1155 | 92.0 | 12.6 | 1011 | 169.9 | 6.0 | 5367 | 1025.7 | 5.2 |  |  |  |  |  |
| 2004 | 498 | 83.1 | 6.0 | 491 | 128.2 | 3.8 | 2527 | 952.1 | 2.7 |  |  |  |  |  |
| 2005 | 506 | 79.1 | 6.4 | 359 | 113.3 | 3.2 | 1651 | 874.2 | 1.9 |  |  |  |  |  |
| 2006 | 548 | 55.6 | 9.8 | 465 | 108.3 | 4.3 | 2229 | 866.8 | 2.6 |  |  |  |  |  |
| 2007 | 886 | 63.4 | 14.0 | 630 | 85.1 | 7.4 | 2995 | 805.7 | 3.7 |  |  |  |  |  |
| 2008 | 591 | 54.0 | 11.0 | 534 | 78.1 | 6.8 | 2284 | 655.3 | 3.5 |  |  |  |  |  |
| 2009 | data unavailable |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.2.8b. Cod in Divisions VIIe-k. Time-series of landings, effort and lpue.

| UK (England and Wales) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Uk Otter trawlers VIIe-k |  |  | Uk Beam trawlers VIIe-k |  |  | Uk Otter trawlers VIIe |  |  |
| Year | Landings | Effort | lpue | Landings | Effort | lpue | Landings | Effort | lpue |
| 1972 | 355 | 117.1 | 3.0 |  |  |  | 80 | 64.6 | 1.2 |
| 1973 | 223 | 118.5 | 1.9 |  |  |  | 58 | 69.5 | 0.8 |
| 1974 | 192 | 91.6 | 2.1 |  |  |  | 55 | 50.1 | 1.1 |
| 1975 | 136 | 100.3 | 1.4 |  |  |  | 38 | 54.7 | 0.7 |
| 1976 | 97 | 88.2 | 1.1 |  |  |  | 32 | 56.1 | 0.6 |
| 1977 | 119 | 88.5 | 1.3 |  |  |  | 78 | 55.4 | 1.4 |
| 1978 | 116 | 83.2 | 1.4 | 6 | 24.7 | 0.3 | 70 | 48.8 | 1.4 |
| 1979 | 130 | 73.5 | 1.8 | 14 | 44.0 | 0.3 | 74 | 49.9 | 1.5 |
| 1980 | 228 | 85.6 | 2.7 | 39 | 76.7 | 0.5 | 84 | 50.0 | 1.7 |
| 1981 | 324 | 104.3 | 3.1 | 63 | 87.6 | 0.7 | 76 | 46.9 | 1.6 |
| 1982 | 362 | 104.7 | 3.5 | 84 | 115.0 | 0.7 | 65 | 38.5 | 1.7 |
| 1983 | 163 | 82.1 | 2.0 | 84 | 135.3 | 0.6 | 73 | 52.6 | 1.4 |
| 1984 | 237 | 86.7 | 2.7 | 129 | 131.5 | 1.0 | 77 | 52.9 | 1.5 |
| 1985 | 249 | 90.3 | 2.8 | 145 | 152.5 | 1.0 | 64 | 57.7 | 1.1 |
| 1986 | 233 | 84.7 | 2.8 | 164 | 135.7 | 1.2 | 80 | 49.5 | 1.6 |
| 1987 | 221 | 84.3 | 2.6 | 246 | 177.1 | 1.4 | 96 | 45.1 | 2.1 |
| 1988 | 270 | 89.1 | 3.0 | 248 | 194.9 | 1.3 | 155 | 53.4 | 2.9 |
| 1989 | 186 | 84.1 | 2.2 | 230 | 198.2 | 1.2 | 105 | 54.7 | 1.9 |
| 1990 | 314 | 99.5 | 3.2 | 307 | 207.6 | 1.5 | 128 | 53.1 | 2.4 |
| 1991 | 243 | 76.7 | 3.2 | 258 | 203.2 | 1.3 | 84 | 40.8 | 2.0 |
| 1992 | 232 | 86.4 | 2.7 | 256 | 196.1 | 1.3 | 81 | 39.9 | 2.0 |
| 1993 | 181 | 61.9 | 2.9 | 220 | 208.4 | 1.1 | 43 | 39.2 | 1.1 |
| 1994 | 79 | 53.7 | 1.5 | 174 | 220.0 | 0.8 | 41 | 38.8 | 1.1 |
| 1995 | 115 | 52.3 | 2.2 | 239 | 243.1 | 1.0 | 55 | 35.5 | 1.5 |
| 1996 | 120 | 60.5 | 2.0 | 303 | 260.8 | 1.2 | 59 | 30.5 | 1.9 |
| 1997 | 149 | 66.7 | 2.2 | 299 | 264.8 | 1.1 | 79 | 33.3 | 2.4 |
| 1998 | 119 | 62.1 | 1.9 | 265 | 254.6 | 1.0 | 62 | 29.8 | 2.1 |
| 1999 | 90 | 98.4 | 0.9 | 257 | 251.4 | 1.0 | 47 | 27.5 | 1.7 |
| 2000 | 111 | 104.1 | 1.1 | 187 | 259.0 | 0.7 | 52 | 30.5 | 1.7 |
| 2001 | 110 | 85.3 | 1.3 | 256 | 272.7 | 0.9 | 59 | 31.9 | 1.8 |
| 2002 | 80 | 83.0 | 1.0 | 130 | 249.5 | 0.5 | 34 | 28.3 | 1.2 |
| 2003 | 58 | 72.3 | 0.8 | 103 | 282.1 | 0.4 | 24 | 25.1 | 1.0 |
| 2004 | 44 | 75.7 | 0.6 | 96 | 273.9 | 0.3 | 15 | 25.6 | 0.6 |
| 2005 | 41 | 76.4 | 0.5 | 102 | 270.3 | 0.4 | 17 | 21.1 | 0.8 |
| 2006 | 55 | 83.3 | 0.7 | 91 | 252.0 | 0.4 | 13 | 21.1 | 0.6 |
| 2007 | 49 | 87.6 | 0.6 | 111 | 239.9 | 0.5 | 22 | 22.4 | 1.0 |
| 2008 | 49 | 71.2 | 0.7 | 71 | 216.9 | 0.3 | 24 | 19.9 | 1.2 |
| 2009 | 27 | 73.8 | 0.4 | 67 | 190.9 | 0.4 | 13 | 21.4 | 0.6 |

Units: landings in tonnes live weight, Effort in 000s hours fished, lpue in $\mathrm{Kg} /$ hour fished.

Table 7.2.8c. Cod in Divisions VIIe-k. Time-series of landings, effort and lpue.

|  | IRELAND |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ir Otter trawlers VIIg |  |  | Ir Beam trawlers VIIg |  |  | Ir Scottish seiners VIIg |  |  | Ir Gillnet VIIg |
| Year | Landings | Effort 1 | lpue | Landings | Effort |  | Landings | Effort | lpue | Landings Effort lpue |
| 1995 | 429.9 | 63.6 | 6.8 | 85.8 | 20.8 | 4.1 | 111.27 | 6.43 | 17.3 | 114.92 |
| 1996 | 569.3 | 60.0 | 9.5 | 112.6 | 26.8 | 4.2 | 164.87 | 9.73 | 16.9 | 338.84 |
| 1997 | 401.9 | 65.1 | 6.2 | 131.6 | 28.3 | 4.7 | 215.24 | 16.13 | 13.3 | 52.81 |
| 1998 | 450.6 | 72.3 | 6.2 | 166.9 | 35.3 | 4.7 | 264.14 | 14.94 | 17.7 | 87.32 |
| 1999 | 300.9 | 51.7 | 5.8 | 190.6 | 40.9 | 4.7 | 64.59 | 8.01 | 8.1 | 211.92 |
| 2000 | 279.4 | 60.6 | 4.6 | 180.7 | 37.0 | 4.9 | 106.04 | 9.90 | 10.7 | 157.03 |
| 2001 | 339.5 | 69.4 | 4.9 | 96.6 | 39.7 | 2.4 | 111.09 | 16.33 | 6.8 | 107.99 |
| 2002 | 213.0 | 77.7 | 2.7 | 57.9 | 31.6 | 1.8 | 70.84 | 20.86 | 3.4 | 34.13 |
| 2003 | 167.4 | 86.8 | 1.9 | 57.1 | 49.3 | 1.2 | 38.07 | 20.91 | 1.8 | 31.17 |
| 2004 | 190.2 | 97.0 | 2.0 | 74.3 | 54.9 | 1.4 | 54.86 | 19.38 | 2.8 | 60.65 |
| 2005 | 294.9 | 124.4 | 2.4 | 118.7 | 49.7 | 2.4 | 66.13 | 14.81 | 4.5 | 77.697 |
| 2006 | 390.0 | 119.2 | 3.3 | 128.6 | 60.5 | 2.1 | 90.98 | 14.79 | 6.2 | 63.73 |
| 2007 | 323.0 | 136.5 | 2.4 | 96.2 | 55.9 | 1.8 | 58.52 | 15.82 | 3.7 | 85.44 |
| 2008 | 349.9 | 125.8 | 2.8 | 85.4 | 37.2 | 2.3 | 55.59 | 11.65 | 4.8 | 86.77 |
| 2009 | 402.6 | 135.2 | 3.0 | 74.1 | 37.9 | 2.0 | 34.51 | 8.15 | 4.2 | 80.88 |

Units: landings in Tonnes live weight, Effort in 000s
hours fished, lpue in $\mathrm{Kg} /$ hour fished

|  | IRELAND |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ir Otter trawlers VIIj |  |  | Ir Beam trawlers VIIj |  |  | Ir Scottish seiners VIIj |  |  | Ir Gillnet VIIj |
| Year | Landings | Effort | lpue | Landings | Effort | lpue | Landings | Effort | lpue | Landings Effort lpue |
| 1995 | 338.5 | 93.7 | 3.6 | 0.1 | 0.2 | 0.2 | 75.52 | 5.26 | 14.4 | 179.57 |
| 1996 | 326.4 | 70.2 | 4.6 | 8.7 | 1.5 | 5.9 | 124.55 | 8.15 | 15.3 | 64.96 |
| 1997 | 352.8 | 83.2 | 4.2 | 3.4 | 1.8 | 1.9 | 115.81 | 10.73 | 10.8 | 45.47 |
| 1998 | 262.3 | 89.6 | 2.9 | 19.2 | 5.2 | 3.7 | 103.37 | 6.61 | 15.6 | 59.13 |
| 1999 | 76.7 | 40.6 | 1.9 | 27.6 | 7.4 | 3.7 | 9.57 | 1.41 | 6.8 | 24.01 |
| 2000 | 95.5 | 64.6 | 1.5 | 21.2 | 6.9 | 3.1 | 23.71 | 3.49 | 6.8 | 13.98 |
| 2001 | 140.4 | 67.7 | 2.1 | 10.4 | 3.0 | 3.5 | 27.95 | 4.42 | 6.3 | 12.69 |
| 2002 | 150.1 | 90.4 | 1.7 | 5.4 | 3.1 | 1.7 | 24.65 | 8.87 | 2.8 | 12.23 |
| 2003 | 78.5 | 111.3 | 0.7 | 8.8 | 9.0 | 1.0 | 14.72 | 9.15 | 1.6 | 6.17 |
| 2004 | 36.1 | 92.0 | 0.4 | 2.5 | 2.2 | 1.2 | 11.57 | 9.18 | 1.3 | 4.21 |
| 2005 | 40.6 | 73.9 | 0.5 | 4.7 | 2.4 | 1.9 | 17.76 | 6.09 | 2.9 | 3.30 |
| 2006 | 42.7 | 65.9 | 0.6 | 2.0 | 1.5 | 1.3 | 15.64 | 5.33 | 2.9 | 7.18 |
| 2007 | 39.0 | 80.5 | 0.5 | 7.8 | 2.4 | 3.3 | 9.83 | 3.51 | 2.8 | 6.50 |
| 2008 | 33.5 | 66.5 | 0.5 | 2.6 | 1.1 | 2.3 | 9.46 | 2.84 | 3.3 | 6.50 |
| 2009 | 26.2 | 72.5 | 0.4 | 4.7 | 2.8 | 1.7 | 8.90 | 3.33 | 2.7 | 7.78 |

Units: landings in Tonnes live weight, Effort in 000s
hours fished, lpue in $\mathrm{Kg} /$ hour fished

Table 7.2.9. List of parameters and data used for the YPR analysis.

| Age group | Selectivity fishing mortality | Selectivity natural mortality | Stock weights | Catch weights | Spawningstock weights | Fraction mature |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.19 | 0.20 | 0.88 | 0.88 | 0.613 | 0.01 |
| 2 | 0.84 | 0.20 | 2.07 | 2.07 | 1.535 | 0.39 |
| 3 | 0.86 | 0.20 | 4.68 | 4.68 | 4.315 | 0.87 |
| 4 | 0.63 | 0.20 | 7.36 | 7.36 | 7.201 | 0.93 |
| 5 | 0.51 | 0.20 | 9.58 | 9.58 | 9.429 | 1.00 |
| 6 | 0.40 | 0.20 | 11.67 | 11.67 | 11.492 | 1.00 |
| 7 | 0.40 | 0.20 | 13.44 | 13.44 | 13.359 | 1.00 |
| 8 | 0.40 | 0.20 | 15.44 | 15.44 | 15.461 | 1.00 |
| 9 | 0.40 | 0.20 | 16.20 | 16.20 | 15.894 | 1.00 |
| 10 | 0.40 | 0.20 | 16.32 | 16.32 | 15.744 | 1.00 |
| Natural mortality: |  |  | 0.20 |  |  |  |
| Proportion of fishing mortality before spawning: |  |  | 0.00 |  |  |  |
| Proportion of natural mortality before spawning: |  |  | 0.00 |  |  |  |



Figure 7.2.1. Irish industry and science survey. Maps of station and priority.







Figure 7.2.2a. Cod in Divisions VIIe-k. 2009 Quarterly or annual length compositions of UK, Irish discards raised using effort ratio for Irish data, from hauls sampled for UK.







Figure 7.2.2b. Cod in Divisions VIIe-k. 2008 Quarterly length compositions of French catches in VIIfgh, selfsampling data from FR-GADOID raised by landings ratio



Figure 7.2.2b. Continued.








Figure 7.2.2c. Cod in Divisions VIIe-k. 2009 Quarterly length composition of French landings and discards available from hauls sampled by observers at sea.


Figure 7.2.2d. Cod in Divisions VIIe-k. 2009 Quarterly length composition of French landings and discards. Self-sampling programme.


Figure 7.2.3. Cod in Divisions VIIe-k. Percentage of landings accounted for by each age class in Celtic Sea cod over the time-series (Data Source: WGCSE 2009).

FR-EVHOE Groundfish Oct-Nov survey in VIIf,g,h,j, numbers per $30 \mathrm{mn} \square$

/HOE Groundfish Oct-Nov survey in VIIf,g,h,j, numbers per 30 mñComparative scatterplots at


FR-EVHOE Groundfish Oct-Nov survey in VIIf,g,h,j, numbers per $30 \mathrm{mn} \square$


FR-EVHOE Groundfish Oct-Nov survey in VIIf,g,h,j, numbers per 30 mn■og cohort abundance


FR-EVHOE Groundfish Oct-Nov survey in VIlf,g,h,j, numbers per $30 \mathrm{mn} \square$



Figure 7.2.4a. Cod in VII e-k. Diagnostics SURBA v3.0 plots for FR-EVHOE survey, age groups 15. Log mean-standardised indices by year and age class, scatter plots, catch curves, and residuals (Single fleet). Bottom right: Raw abundance indices (number of cods caught per $\mathbf{3 0} \mathbf{~ m i n}$ tow) from the FR-EVHOE survey.


Figure 7.2.4b. Cod in VII e-k. Diagnostics SURBA v3.0 plots for IR-GFS7gj survey, age groups 15. Log mean standardised indices by year and age class, scatter plots, catch curves, and residuals. (Single fleet).


UK-WCGFS West Coast March survey, effort in $m n$ towed, numbers *10**2, final survey in 2004


Figure 7.2.4c. Cod in VII e-k. Diagnostics of WGCSE 2009 SURBA v3.0 plots for UKWCGFS survey, age groups 1-5. Log mean standardised indices by year and age class, scatter plots, catch curves, and residuals. (Single fleet).


Mean total mortality


UK-WCGFS WGCSE 2009

Figure 7.2.4d. Cod in VII e-k. Trends of relative mean Z. SURBA v3.0 plots for the 3 surveys used separately.


Figure 7.2.4e. Comparative trends of age 1 index of FR-EVHOE survey and recruitment estimates from a Separable VPA run.



Figure 7.2.5a. Cod in Divisions VIIe-k. Trends of lpues and effort. French Gadoid trawlers and French Nephrops trawlers in VIIfgh.


Figure 7.2.5a. Continued. Cod in Divisions VIIe-k. Trends of lpues and effort. French otter trawlers in VIIe-k (including Gadoid trawlers and Nephrops trawlers in VIIfgh) and French otter trawlers in VIIe.


Figure 7.2.5a. Continued. Cod in Divisions VIIe-k. Trends of lpues and effort. UK otter trawlers in VIIe-k and VIIe, UK beam trawlers in VIIe-k.



Figure 7.2.5b. Cod in Divisions VIIe-k. Trends of lpues and effort. Irish otter trawlers in VIIg and VIIj, Irish beam trawlers in VIIg and VIIj.



Figure 7.2.5b. Cod in Divisions VIIe-k. Trends of lpues and effort. Irish Scottish seiners in VIIg and VIIj.


Figure 7.2.6. Cod in VII e-k. Distribution of landings by otter trawlers in the TAC area.


Figure 7.2.7. Cod in VII e-k. Distribution of effort by French and Irish otter trawlers in the TAC area.


Figure 7.2.8. Cod in VII e-k. Distribution of lpues by French and Irish otter trawlers in the TAC area.


Figure 7.2.9. Exploratory catch curve analysis on 1971-2009 and 2000-2009 Celtic Sea cod catch-atage data.


Figure 7.2.10. Cod in VIIe-k. Exploratory yield and SSB per recruit.



Figure 7.2.11. Cod in VIIe-k. Sensibility analysis on $F_{\text {max }}$ and $F_{0.1}$ to a change in natural mortality.

### 7.3 Cod in Divisions VIIb, c

## Type of assessment: No assessment

The nominal landings are given in Table 7.3.1.
Table 7.3.1. Landings $(\mathbf{t})$ of cod in Division VIIb,c for 1995-2009 as officially reported to ICES.

| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | 91 | 115 | 71 | 44 | .$^{1}$ | 46 | 38 | 54 |
| Germany | - | - | 3 | - | - | - | - | - |
| Ireland | 282 | 353 | 177 | 234 | 154 | 141 | 107 | 59 |
| Netherlands | - | - | - | - | - | - | + | - |
| Norway | 3 | 1 | 6 |  | 11 | +* | 1 | 5 |
| Spain | 6 | 3 |  | 6 | 2 | 3 | 1 | 1 |
| UK(E/W/NI) | 25 | 35 | 37 | 25 | 4 | 4 | 2 | 1 |
| UK(Scotland) | 66 | 12 | 7 | 9 | 1 | - |  | 1 |
| UK |  |  |  |  |  |  |  |  |
| Total | 473 | 519 | 301 | 318 | 172 | 194 | 150 | 122 |
| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |
| France | 33 | 13 | 13 | 10 | 18 | 14 |  |  |
| Germany |  |  |  |  |  |  |  |  |
| Ireland | 59 | 60 | 32 | 16 | 11 | 18 | 17 |  |
| Netherlands | 1 |  |  |  |  |  |  |  |
| Norway |  |  |  | 1 | 1 |  |  |  |
| Spain |  |  |  |  |  |  |  |  |
| UK(E/W/NI) | 8 |  | 0 | 1 | 2 | 1 |  |  |
| UK(Scotland) | 1 | 10 |  | 0 |  |  |  |  |
| UK |  |  |  |  |  |  | 0 |  |
| Total | 102 | 83 | 45 | 28 | 32 | 33 | 17 |  |

### 7.4 Haddock in Divisions VIIb-k

Type of assessment in 2010
Update.
ICES advice applicable to 2009 and 2010
"Last year's advice remains unchanged: Future catches and SSB will be highly dependent on the strength of incoming year classes and their discard mortality. No strong recruitment has been observed since 2002 and estimated recruitment for 2006 is the lowest since 1997. In this context the stock should be managed by ensuring that fishing effort is not allowed to increase."

### 7.4.1 General

## Stock description and management units

The basis for the stock assessment Area VIIb-k is described in detail in the Stock Annex. The TAC for haddock is set for all of Subarea VIIb-k, VIII, IX and X, which does not correspond to the stock assessment area (VIIb-k). However, official international landings from VIII, IX and X have been less than $2 \%$ of all landings in the TAC area in most years since 1973.


Red Boxes-TAC/Management Areas Blue Shading- Assessment Area.

Management applicable to 2009 and 2010

TAC table 2009

| Species: Haddock <br> Melanogrammus aegkfinus |  | Zone: | VIIb-k, VIII, IX and X; EC waters of CECAF 34.1.1 <br> (HAD/7X7A34) |
| :---: | :---: | :---: | :---: |
| Belgium | 129 |  |  |
| France | 7719 |  |  |
| Ireland | 2573 |  |  |
| United Kingdom | 1158 |  |  |
| EC | 11579 |  |  |
| TAC | 11579 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TAC table 2010

| Species: | Haddock <br> Melanogrammus aeglefinus | Zone: | VIIb-k, VIII, IX and X; EU waters of CECAF 34.1.1 <br> (HAD/7X7A34) |
| :--- | ---: | :--- | :--- |
| Belgium | 129 |  |  |
| France | 7719 |  |  |
| Ireland | 2573 |  |  |
| United Kingdom | 1158 |  |  |
| EU | 11579 | Analytical TAC |  |
| TAC | 11579 |  |  |

Since 2009, a separate TAC is set for VIIa haddock, previously a separate allocation for VIIa existed within the TAC for VII, VIII, IX and X.

## Fishery in 2009

The official landings reported to ICES and Working Group estimates of the landings and discards are given in Table 7.4.1. France did not submit official landings and the Irish official landings ( 1794 t ) appear to be incorrect and were considerably lower than those used by the Working Group (2966 t). France, the UK and Ireland provided minor revisions to the landings figures for 2008. The 2008 landings figure of 7013 t was revised to 7049 t . The 2009 landings were estimated by the WG to be 10028 t .

Before 2002, the TAC was well in excess of the landings in the TAC area. During 2002, 2003 and 2004 the TAC was reduced to less than 10000 t and it appeared to be restrictive. (WGSSDS05 provided some qualitative evidence that misreporting was now a problem). During 2005-2008 the TAC was between 11520 t and 11579 t and the landings in the TAC area were less than $70 \%$ of the TAC. In 2009 the total landings (WG estimates) are still lower than the TAC but the quota appeared to be restrictive for Ireland and Belgium (WG landings of 2966 t and 131 t respectively) but not for France and the UK (WG landings of 6230 t and 703 t respectively).

### 7.4.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1.

## Numbers-at-length

Length compositions of landings were available for haddock landed into Ireland, France and the UK in 2009 (Table 7.4.2; Figure 7.4.1). Length distributions of the various fleets are quite similar.

Discard length distributions for 2008 are shown in Figure 7.4.2. The figure shows that there are considerable numbers of small $(<20 \mathrm{~cm})$ haddock in the Irish discards in VIIb and VIIg and in the UK discards. These fish first appeared in the third quarter of 2009 and are likely to be 0-group haddock. The figure also shows that there appears to be considerable discarding of haddock over the minimum landing size by the French gadoid and benthic fleet and the UK fleets.

## Discard numbers-at-age

Irish otter trawl discard data were raised to the national level using the number of trips as auxiliary variable as described in the Stock Annex. The numbers of OTB discard trips by year and métier are given in Table 7.4.3a and the total number of OTB trips is given in Table 7.4.3b. Irish discard data from VIIgj were used to estimate international discards by using the ratio of the international effort in VIIe-k to the Irish effort in VIIgj (Table 7.4.3c). French effort data were not available and the average 2006-2008 effort was assumed for 2009.

Figure 7.4.3a shows the Irish discard numbers-at-age and the discard numbers-at-age raised to international levels. Figure 7.4.3b shows the proportions-at-age that are discarded; over the last 10 years $88 \%$ of 1-year-olds have been discarded, $49 \%$ of 2-yearolds and $16 \%$ of 3 -year-olds have been discarded. By number, $66 \%$ of the total catch was discarded, by weight $41 \%$ was discarded (average last 10 years).

## Landings numbers-at-age

Landings numbers-at-age were raised using the procedure described in the Stock Annex with the exception of the French data. Due to problems with the French logbooks database, the landings data were not available by quarter and métier. To address this, the annual catch was allocated to quarters using the mean proportion by quarter over the period 2006-2008, which appeared to be reasonably stable. Secondly the sample length distributions within each quarter were assumed to be representative of the landings of each métier.

Landings numbers-at-age are given in Table 7.4.4a, discard numbers-at-age are given in Table 7.4.4b and catch numbers-at-age in Table 7.4.4c. Despite uncertainty about the quality of the discard data, it is possible to track strong year classes in both the discards and the landings-at-age matrices. Figure 7.4.4 shows the age compositions of the catches, the figure shows that discards account for a large proportion of the catch numbers up to age 3 .

Mean landings weights-at-age are given in Table 7.4.5a, catch weights-at-age are given in Table 7.4.5b and stock weights are given in Table 7.4.5c.

## Biological

The assumptions of natural mortality and maturity are described in the Stock Annex. The maturity ogive used in the assessment is knife-edged at age 2. Irish Q1 survey
data from 2004-2009 in VIIbgj (WD 3) suggested a similar maturity ogive for females but also indicated that a significant number of males mature before the age of two.

## Surveys and commercial tuning fleets

The surveys are described in the Stock Annex. Available survey indices and tuning fleet data are given in Table 7.4.6. Survey data tuning-series were made available by Ireland, the UK, and France. Commercial tuning fleets were made available by Ireland; the French tuning fleet was not available for 2009 due to problems with the French logbooks database.

The standardised indices are given by year in Figure 7.4.5a and by cohort in Figure 7.4.5b. In addition to the indices that were used in the assessment, the Irish Groundfish Survey (IGFS-IBTS-EA-4Q) indices in VIIb and VIIj are shown. The EVHOE-IBTS-EA-4Q survey is noisy and has a strong year-effect in 2000 but on further analysis did show patterns that were consistent with other surveys, particularly for ages 0 , 2,3 and 4.

Figure 7.4 .6 shows the standardised recruitment (age 0) indices for the EVHOE, SAGFS and IGFS VIIj surveys (the latter is not used in the assessment). All surveys indicate that the 2009 recruitment is the highest in the time-series. The EVHOE survey estimates the 2009 cohort to be $41 / 2$ times as high as average recruitment over the rest of the survey time-series, the SAGFS estimates it to be nearly seven times higher than average and the IR-GFS VIIj index is nearly eight times higher than average.

## Commercial Ipue

Effort and lpue data are given in Table 7.4.7 and Figure 7.4.7.

## Other relevant data

No specific issues were raised by the industry on VIIb-k haddock.

### 7.4.3 Historical stock development

Model used: eXtended Survival Analysis (XSA)
Software used: FLR, VPA95
Exploratory data analysis and the assessment were carried out using FLR under R version 2.8.1 with packages FLCore 2.2, FLAssess 1.99-102, FLXSA 1.99-100 and FLEDA 2.0. The final assessment was also run using the Lowestoft VPA95 software.

## Data screening

The general approach to data screening and analysis was followed in addition to the data exploration tools available in the FLR package FLEDA. The results of the data screening are available in the folder 'Data $\backslash$ Stock $\backslash$ had- $7 \mathrm{~b}-\mathrm{k} \backslash$ Exploratory runs' on SharePoint.

One particular exploratory assessment will be highlighted here: The sensitivity of the assessment to discard data was investigated by setting the catch numbers-at-ages 0,1 and 2 to zero. For these ages the total mortality is then assumed to be equal to the natural mortality. This exploratory assessment showed virtually identical estimates for Fbar 3-5 while SSB and recruitment showed very similar trends. This suggests that the trends-only assessment is not excessively sensitive to inclusion of discard num-bers-at-age.

## Final update assessment

The final assessment was run with the same settings as last year. The only difference is that no data were available for the French commercial tuning fleet (FR7fgGAD) in 2009.

Input data types and characteristics:

| Type | Name | Year range | Age range |
| :--- | :--- | ---: | :---: |
| Caton | Catch in tonnes | $1993-2009$ | $0-8+$ |
| Canum | Catch-at-age in numbers | $1993-2009$ | $0-8+$ |
| Weca | Weight-at-age in the catch | $1993-2009$ | $0-8+$ |
| West | Weight-at-age at spawning time. | $1993-2009$ | $0-8+$ |
| Mprop | Proportion of M before spawning | $1993-2009$ | $0-8+$ |
| Fprop | Proportion of F before spawning | $1993-2009$ | $0-8+$ |
| Matprop | Proportion mature-at-age | $1993-2009$ | $0-8+$ |
| Natmor | Natural mortality | $1993-2009$ | $0-8+$ |

A plusgroup of 8+ was used. Age group 0 was included in the assessment data to allow inclusion of 0 -group indices in the XSA runs. However, catch numbers-at-age 0 were set to zero to avoid spurious F-shrinkage effects at this age.

Model Options:

| Option | Setting |  |
| :--- | :--- | :---: |
| Ages catch dep stock size | None |  |
| Q plateau |  | 4 |
| Taper | No |  |
| F shrinkage SE |  | 1.5 |
| F shrinkage year range | 5 |  |
| F shrinkage age range |  | 3 |
| Fleet SE threshold | No | 0.3 |
| Prior weights |  |  |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :---: | :---: |
| Survey | UK7efghjWCS | $1996-2004$ | Not used |
| Survey | FR7fghjEVHOE | $1997-$ present | $0-5$ |
| Survey | IR7bjWCGFS | $1999-2002$ | Not used |
| Survey | IR7gSAGFS | $1999-$ present |  |
| Survey | IR7bIGFS | $2003-$ present | Not used |
| Survey | IR7jIGFS | $2003-$ present | Not used |
| Commercial | IR7bjOTB | $1995-$ present | $2-7$ |
| Commercial | FR7fghGAD | $2002-2008$ | $2-6$ |

The XSA diagnostics are given in Table 7.4.8. The estimated fishing mortality is quite variable. The catchability regressions and residuals are given in Figure 7.4.8, the residuals are relatively large and some year effects are apparent. The catchability regression for age 7 is very tight, suggesting that the model adjusts the population
numbers at that age to the tuning data. Increasing the fleet SE threshold can prevent this; however the assessment results do not change noticeably when the SE is increased therefore last year's settings were not changed.

The weighting applied to the terminal survivor estimates is shown in Figure 7.4.9. The 2009 cohort takes equal weight from the two surveys. The French Gadoid fleet gets relatively little weight because no 2009 data were available. F-shrinkage does not account for much of the weighting in any of the cohorts.

The retrospective analysis was run back to 2002. The results are shown in Figure 7.4.10. Recruitment in 2002 was initially estimated to be very high (similar to 2009) but was subsequently reduced as more data became available. The survey tuning fleets consist of short time-series which might account for a larger retrospective changes in the estimated recruitment early in the time-series than in more recent years. Fbar was consistently revised upwards up to 2005 but has not shown a retrospective pattern since.

## Comparison with previous assessments

The XSA settings have not changed since 2007 and revisions to previous years' data were minor. This year's assessment did not have any 2009 data for the French gadoid tuning fleet. The last year's estimates of Fbar, SSB, and recruitment have not changed significantly.

## State of the stock

The state of the stock is not precisely known. However SSB has shown an increasing trend over the time-series.

The stock summary is given in Table 7.4.9 and Figure 7.4.11. Following good recruitment in 1999, 2001 and 2002 the SSB and catch increased, however, due to high discarding, the landings (and TAC) did not increase in line with the increased stock levels. Recruitment also has been relatively high in 2007-2008 and the catches in 2009 have increased, but most of these catches were discarded. Recruitment of the 2009 year class appears to be exceptionally good, however under the current discarding pattern it is likely that many of these fish will be discarded before they are of marketable size.

### 7.4.4 Short-term projections

Short-term projections are presented here for reference only; they are not considered reliable for the following reasons:

- It appears that the 2009 cohort is exceptionally strong; however the accuracy of the recruitment estimate is unknown and will have a very large influence on the short-term forecast. The last time that similar recruitment was retrospectively estimated was in 2001 when the survey tuning fleets series were still very short.
- Recruitment of haddock is characterised by sporadic events, therefore the use of geometric mean recruitment (1993-2007) for 2010-2012 is questionable.

Short-term projections were performed using MFDP1a software.
Recruitment for 2010-2012 was estimated at 36586 (GM 93-07; thousands). Three year averages were used for F and weights-at-age. Input data for the short-term fore-
cast are given in Table 7.4.10. Landings and discard numbers and weights were supplied separately. Table 7.4 .11 gives the management options. The short-term forecasts are highly influenced by the strong 2009 year class. Estimates of the relative contribution of recent year classes to the 2011 landings and 2012 SSB are shown in Table 7.4.12. The high recruitment in 2009 accounts for $85 \%$ of the projected landings in 2011 and for $67 \%$ of the SSB in 2012

Conclusion: the short-term projections and management options are highly dependent on the accuracy of the estimated size of the 2009 cohort and are not considered to be reliable.

### 7.4.5 MSY evaluation

There are a number of major points that should be considered when interpreting the MSY analysis:

- Haddock stocks are characterised by extreme recruitment events; recruitment modelled from a stock-recruitment (SR) relationship is therefore only a useful concept in the long term. Additionally, the time-series is quite short and there is little information to inform the SR model.
- The yield in this analysis refers to landings only and is based on the current selectivity pattern. If the selectivity is improved, the MSY reference points will change.
- The assessment is accepted for trends only and that F reference points should therefore only be interpreted in a broad sense, i.e. current F appears to be well above any candidate F target reference point.

MSY estimates were evaluated using the srmsymc ADMB package.
Yield and $\mathrm{F}_{\mathrm{bar}}$ refer to landings only. The potential yield of the discards is not taken into account nor is the mortality due to discarding. $\mathrm{F}_{\text {msy }}$ in this context is therefore the F at which the landings are maximised while accepting a continuation of current discard mortality. Figure 7.4.12 illustrates the estimates of the catch, landings and discard components of F and the selectivity patterns estimated from them.

The R-package FLBRP was used to investigate the sensitivity to the averaging options to YPR estimates. Figure 7.4 .13 shows that the F target reference points did not appear to be very sensitive to the number of years used to estimate average $F$ and average weights-at-age so the default three-year averaging period was maintained.

Figure 7.4.14 shows three stock-recruitment relationships fitted by srmsymc. There is little information in the data to inform the shape stock-recruit relationship and no single model provided the best mathematical fit (Table 7.4.13 provides AIC values for all three models). Sporadic exceptional recruitment is a feature of many haddock stocks; therefore the 2009 recruitment was retained although it had a large influence on the fit of the SR models.

Figures 7.4.15 to 7.4.17 show box plots of $\mathrm{F}_{\text {msy }}$ and $\mathrm{F}_{\text {crash }}$ as well as $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim. }}$ Table 7.4.13 summarises the MSY evaluation. The BH and hockey stick models estimate current F to be above any of the candidate F reference points but the Ricker model suggests that $\mathrm{F}_{\text {msy }}$ is above current F . The Ricker model assumes impaired recruitment at high stock levels and therefore arrives at higher $\mathrm{F}_{\text {msy }}$ than the other models. This level of $\mathrm{F}_{\text {msy }}$ overlaps with the confidence limits of $\mathrm{F}_{\text {crash }}$ and is therefore not an appropriate F target. Therefore the Ricker model was rejected. The BH and hockey stick models both result in Fmsy estimates that are below current F.

Conclusion: The stock-recruit relationship of haddock is not well captured by any of the models and the underlying data do not support the provision of absolute estimates of $F_{\text {msy. }}$. However it is likely that current $F$ is above $F_{\text {msy }}$.

### 7.4.6 Biological reference points

Precautionary approach reference points
It is not possible to derive precautionary reference points for this stock from the short time-series of information available.

### 7.4.7 Management plans

No management plan for VIIbk haddock has been agreed or proposed.

### 7.4.8 Uncertainties and bias in assessment and forecast

## Landings

The sampling levels of landings for countries supplying data for 2009 are given in Table 2.1. Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches, although the assessment is contingent on the accuracy of the landings statistics. The French landings data were not fully available and the catch numbers-at-age data were raised using an ad hoc approach as described in 7.4.2. The aggregated landings data were fully available; however, the lack of quarterly and métier data means that the weighting of the sampling data might be inaccurate. It is difficult to estimate how this will have affected the assessment.

## Discards

Sampling levels for discarding are low, resulting in a high level of variability. Discards account for more than half of the catch weight in some years and it is therefore very important that they are taken into account. France and the UK have collected discard data in recent years and WD3 (WGCSE09) provides a comparison of the French and Irish discarding data. The method of raising discards data used here has a few undesirable features. Firstly, the data is only from the Irish OTB fleet and this fleet is not the largest component of the overall fishery, although otter trawlers in general do account for most of the haddock VIIb-k fishery. Secondly, a small number of discard trips was raised to account for a very large component of the catch data. The level of variability is unknown but likely to be high. Finally, raising data to international levels using effort, assumes that discard rates per hour trawled of the Irish OTB fleet are similar to all other fleets. It is known that the lpue of the French gadoid fleet is much higher than that of the other fleets so it is likely that the cpue is different as well. The French fleets mainly use $>=100 \mathrm{~mm}$ mesh size while the Irish fleets use both 80 mm and 100 mm mesh. Nevertheless, the discard-at-age matrix did allow tracking of cohorts and there does seem to be merit in including them in the assessment, despite the reservations outlined above. The sensitivity of the assessment to discard data was investigated during the data exploration by running an assessment with the same settings as the final assessment but excluding the age classes for which significant discarding takes place. The results suggest that the current trends-only assessment is not particularly sensitive to inclusion of discard numbers-at-age.

## Surveys

None of the available surveys cover the full assessment area. The EVHOE survey covers the southern end of the area (VIIh+f and the southern part of VIIg+j) while the SAGFS only covers VIIj. The IGFS in VIIb and VIIj is not presently included but should be considered at the next Benchmark.

## Forecast

The short-term forecast is heavily dependent on the accuracy of the estimated size of the 2009 year class. There is insufficient information to predict whether this estimate is likely to be revised downwards in future.

### 7.4.9 Recommendation for next Benchmark

## Review Group comments

Comments from the Review Group were addressed as far as possible without performing a benchmark assessment. The Stock Annex was amended to take some of the Review Group comments into account.

## Recommendations for future work

It is unlikely that the precision and accuracy of the historical discard data can be improved significantly by further analysis and until the time-series of international discards is long enough and of sufficient quality, no benchmark assessment will be proposed. However the following issues can be explored in preparation for any future benchmark.

- Methods of including the French discard data into the assessment need to be investigated and bias in the historical discard (before French data became available) data needs to be investigated and addressed. A first step towards this goal is presented in WGCSE WD3, which concludes that the Irish fleets catch and discard more 1-year-old haddock than the French fleets. The current procedure of raising the Irish discard numbers-at-age is therefore likely to overestimate the number of 1-year-olds.
- The two survey tuning fleets (EVHOE and SAGFS) show very good agreement on the trends in the 0-group (Figure 7.4.6). The new Irish Groundfish Survey in VIIb and VIIj (IRGFS; not used in the analysis) generally agrees with the other surveys. It is believed that a significant amount of recruitment takes place in VIIb and the north of VIIj, these divisions are not covered by the EVHOE or SAGFS indices; therefore it would be worth considering including the IRGFS index at the next Benchmark Assessment.
- EVHOE tuning fleet data from 1997 to 2000 are based on Irish survey AgeLength Keys. The time-series is now sufficiently long to omit these years.
- Commercial tuning fleets might be improved by selecting a subset of vessels that have a consistent spatial and temporal effort and catch composition over a significant period of the time-series. This would require a detailed analysis of vessel behaviour.
- The Review Group suggested that "a model that allows for catch by multiple fleets should be developed to account for differences between countries and gear types. The RG suggests that in order to account for numerous fleets a forward projection, statistical catch-at-age model should be considered in the next benchmark assessment, because it may be a more appropriate method than the XSA model due


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to its increased flexibility. In addition, such a model will allow for error in catch-at-age, which is important for a fishery with such high and uncertain levels of discards. It might also be possible under such a framework to more readily and easily include all the surveys, even those for which sampling has been discontinued."


### 7.4.10 Management considerations

Management by TAC is inappropriate for this stock because landings, but not catches, are controlled.

Discarding is a serious problem for this stock; over the last 10 years $66 \%$ of the catch has been discarded ( $41 \%$ by weight). The discard rate of one-year-olds was $88 \%$; two-year-olds $49 \%$ and three-year-olds $16 \%$ by number.

An increase in mesh size to reduce discarding will be beneficial to this stock and could increase the yield considerably. Reduced selectivity on younger ages would reduce discarding and would promote stock increase when strong year classes occur. In Celtic Sea fisheries, some fleets are using 80 mm mesh to target Nephrops, 90 mm mesh in mixed fisheries and 100 mm to target gadoids and other species. Recent gear trials have shown that square mesh panels can significantly reduce discards of undersized haddock when using Nephrops gear (BIM, 2009). WGSSDS 08 has pointed out that the selection L50 for 90 mm mesh for haddock in VIIg is 19 cm , which is well below the MLS of 30 mm . In order to minimise discards, a square mesh panel of at least 120 mm should be introduced for the Nephrops fleet and a minimum mesh size of at least 100 mm with a square mesh panel of at least 110 mm for all other fleets.

The TAC has not been restrictive in recent years but in 2009 the national quota of Ireland and Belgium appeared to have become restrictive. The catches are likely to increase as the 2009 cohort enters the fishery; a restrictive TAC is likely to result in high-grading in addition to discarding of fish below MLS.

### 7.4.11 References

BIM. 2009. Summary report of Gear Trials to Support Ireland's Submission under Articles 11 \& 13 of Reg. 1342/2008. Nephrops Fisheries VIIa \& VIIb-k. Project 09.SM.T1.01. Bord Iascaigh Mhara (BIM) May 2009.

Table 7.4.1. Landings (t) of haddock in VIIb-k, officially reported to ICES and the landings used by the Working Group.

|  | Official landings |  |  |  |  |  | Un | Used by WG |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Belgium | France | Ireland | UK | Others | Total | allocated | Landings | Discards | Catch |
| 1984 | 0 | 3328 | 646 | 403 | 549 | 4926 |  |  |  |  |
| 1985 | 4 | 2438 | 794 | 175 | 565 | 3976 |  |  |  |  |
| 1986 | 6 | 2279 | 317 | 245 | 86 | 2933 |  |  |  |  |
| 1987 | 12 | 2380 | 314 | 273 | 0 | 2979 |  |  |  |  |
| 1988 | 64 | 3275 | 275 | 409 | 0 | 4023 |  |  |  |  |
| 1989 | 117 | 3412 | 323 | 295 | 27 | 4174 |  |  |  |  |
| 1990 | 22 | 2110 | 461 | 318 | 31 | 2942 |  |  |  |  |
| 1991 | 18 | 1508 | 1020 | 250 | 97 | 2893 |  |  |  |  |
| 1992 | 21 | 1461 | 1073 | 306 | 26 | 2887 |  |  |  |  |
| 1993 | 51 | 1839 | 1262 | 256 | 0 | 3408 | -60 | 3348 | 1193** | 4541 |
| 1994 | 123 | 2788 | 908 | 240 | 17 | 4076 | 55 | 4131 | 1193** | 5324 |
| 1995 | 189 | 2964 | 966 | 266 | 83 | 4468 | 2 | 4470 | 472 | 4942 |
| 1996 | 133 | 4527 | 1468 | 439 | 86 | 6653 | 103 | 6756 | 1403 | 8159 |
| 1997 | 246 | 6581 | 2789 | 569 | 85 | 10270 | 557 | 10827 | 2120 | 12947 |
| 1998 | 142 | 3674 | 2788 | 444 | 312 | 7360 | 308 | 7668 | 356 | 8025 |
| 1999 | 51 | 2725 | 2034 | 278 | 159 | 5247 | -365 | 4882 | 625 | 5507 |
| 2000 | 90 | 3088 | 3066 | 289 | 123 | 6656 | 755 | 7411 | 7057 | 14468 |
| 2001 | 165 | 4842 | 3608 | 422 | 665 | 9702 | -1070 | 8632 | 1952 | 10584 |
| 2002 | 132 | 4348 | 2188 | 315 | 106 | 7089 | -686 | 6403 | 7468 | 13871 |
| 2003 | 118 | 5781 | 1867 | 393 | 82 | 8241 | -95 | 8146 | 8221 | 16367 |
| 2004 | 136 | 6130 | 1715 | 313 | 159 | 8453 | 128 | 8581 | 5371 | 13952 |
| 2005 | 167 | 4174 | 2037 | 292 | 197 | 6867 | -219 | 6648 | 2563 | 9212 |
| 2006 | 99 | 3190 | 1875 | 274 | 209 | 5647 | -264 | 5383 | 2092 | 7474 |
| 2007 | 119 | 4142 | 1930 | 386 | 52 | 6629 | -119 | 6510 | 3252 | 9762 |
| 2008 | 108 | 3639 | 1800 | 566 | 121 | 6234 | 815 | 7049 | 9302 | 16350 |
| 2009* | 131 | - | 1794 | 715 | 1 | 2641 | 7387 | 10028 | 7095 | 17123 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  | Landings us | d by WG |  |  |  |  |
|  |  | Year | Belgium | France | Ireland | UK | Others | Total |  |  |
|  |  | 2002 | 134 | 3878 | 2070 | 301 | 21 | 6403 |  |  |
|  |  | 2003 | 116 | 5960 | 1667 | 362 | 41 | 8146 |  |  |
|  |  | 2004 | 137 | 6336 | 1732 | 303 | 73 | 8581 |  |  |
|  |  | 2005 | 165 | 4096 | 1991 | 282 | 20 | 6555 |  |  |
|  |  | 2006 | 98 | 3151 | 1857 | 262 | 14 | 5383 |  |  |
|  |  | 2007 | 118 | 4073 | 1925 | 383 | 10 | 6510 |  |  |
|  |  | 2008 | 109 | 4587 | 1794 | 545 | 14 | 7049 |  |  |
|  |  | 2009 | 128 | 6230 | 2966 | 703 | 2 | 10028 |  |  |

Table 7.4.2. Length frequency distributions (‘000) of the landings of haddock in VIIb-k in 2009. FR GAD is the French gadoid fleet, IRL OTB is the Irish otter trawl fleet, UK trawl includes all trawl gears except beam trawl.

|  | FR GAD VIIfgh | $\begin{aligned} & \hline \text { IRL OTB } \\ & \text { VIIb } \end{aligned}$ | $\begin{aligned} & \hline \text { IRL OTB } \\ & \text { VIIg } \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { IRL OTB } \\ & \text { VIIj } \\ & \hline \end{aligned}$ | UK Trawl VIle-k | UK Beam VIIe-k |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | $\begin{gathered} \hline \text { Landings } \\ 4490 \\ \hline \end{gathered}$ | Landings 252 | $\begin{gathered} \hline \text { Landings } \\ 1329 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Landings } \\ 575 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Landings } \\ 579 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Landings } \\ 109 \\ \hline \end{gathered}$ |
| 24 | 0.0 | 0.1 | 0.0 | 0.3 | 0.0 | 0.0 |
| 25 | 0.0 | 0.3 | 0.0 | 0.9 | 0.0 | 0.0 |
| 26 | 0.0 | 0.2 | 0.0 | 6.4 | 0.0 | 0.0 |
| 27 | 1.3 | 1.2 | 0.2 | 21.8 | 0.0 | 0.0 |
| 28 | 47.0 | 3.8 | 3.0 | 30.7 | 0.0 | 0.0 |
| 29 | 150.7 | 6.6 | 6.4 | 47.4 | 0.6 | 0.4 |
| 30 | 240.4 | 12.8 | 15.3 | 48.6 | 0.6 | 1.0 |
| 31 | 266.3 | 15.0 | 30.2 | 41.0 | 3.6 | 3.0 |
| 32 | 353.7 | 17.1 | 67.3 | 49.0 | 18.1 | 3.8 |
| 33 | 455.6 | 16.8 | 90.5 | 44.0 | 39.8 | 6.5 |
| 34 | 582.6 | 18.7 | 119.0 | 56.8 | 71.2 | 7.1 |
| 35 | 597.8 | 21.4 | 150.1 | 62.7 | 55.4 | 9.0 |
| 36 | 548.7 | 26.9 | 156.5 | 73.9 | 60.2 | 10.4 |
| 37 | 497.5 | 26.4 | 160.1 | 68.5 | 68.9 | 7.9 |
| 38 | 410.5 | 27.0 | 133.1 | 60.9 | 60.1 | 10.3 |
| 39 | 417.4 | 24.0 | 119.2 | 64.5 | 57.3 | 7.8 |
| 40 | 414.9 | 21.7 | 115.5 | 45.4 | 71.9 | 6.8 |
| 41 | 278.0 | 19.4 | 91.1 | 35.2 | 66.8 | 7.1 |
| 42 | 227.2 | 19.7 | 86.0 | 28.3 | 47.4 | 4.8 |
| 43 | 164.3 | 14.5 | 70.1 | 26.2 | 43.4 | 4.3 |
| 44 | 148.7 | 12.1 | 68.5 | 21.5 | 40.1 | 4.2 |
| 45 | 202.2 | 8.9 | 51.2 | 18.2 | 30.0 | 4.0 |
| 46 | 144.7 | 6.9 | 47.3 | 16.7 | 19.7 | 3.4 |
| 47 | 130.1 | 7.0 | 34.8 | 14.3 | 17.8 | 3.7 |
| 48 | 70.6 | 6.9 | 36.3 | 13.7 | 7.0 | 3.5 |
| 49 | 72.4 | 3.7 | 23.7 | 10.1 | 6.4 | 2.6 |
| 50 | 63.0 | 3.6 | 27.8 | 8.1 | 10.7 | 2.2 |
| 51 | 61.4 | 3.3 | 21.1 | 6.8 | 13.1 | 2.0 |
| 52 | 61.2 | 2.7 | 11.0 | 8.6 | 7.3 | 1.8 |
| 53 | 37.6 | 3.5 | 12.6 | 4.7 | 4.2 | 1.4 |
| 54 | 39.1 | 2.7 | 14.6 | 2.9 | 0.8 | 1.6 |
| 55 | 32.5 | 2.8 | 12.1 | 4.2 | 0.0 | 1.6 |
| 56 | 16.1 | 1.9 | 8.9 | 2.0 | 0.0 | 1.3 |
| 57 | 15.9 | 2.1 | 6.4 | 3.3 | 3.1 | 0.9 |
| 58 | 13.3 | 2.2 | 5.4 | 2.6 | 0.8 | 0.9 |
| 59 | 22.0 | 1.8 | 5.9 | 1.1 | 1.0 | 0.6 |
| 60 | 10.3 | 1.1 | 5.6 | 1.4 | 0.0 | 0.6 |
| 61 | 8.3 | 0.9 | 4.5 | 0.5 | 0.0 | 0.4 |
| 62 | 7.0 | 1.0 | 3.3 | 0.9 | 0.0 | 0.3 |
| 63 | 11.6 | 1.2 | 3.2 | 0.7 | 0.0 | 0.3 |
| 64 | 5.3 | 0.7 | 2.0 | 0.7 | 0.6 | 0.4 |
| 65 | 7.4 | 0.4 | 1.6 | 0.4 | 0.0 | 0.4 |
| 66 | 3.3 | 0.6 | 1.7 | 0.3 | 0.0 | 0.1 |
| 67 | 2.0 | 0.2 | 1.4 | 0.6 | 1.0 | 0.3 |
| 68 | 2.4 | 0.4 | 1.0 | 0.3 | 0.0 | 0.3 |
| 69 | 4.7 | 0.1 | 0.2 | 0.8 | 0.0 | 0.0 |
| 70 | 2.3 | 0.1 | 0.3 | 0.5 | 0.0 | 0.0 |
| 71 | 1.2 | 0.2 | 0.5 | 0.3 | 0.0 | 0.0 |
| 72 | 1.6 | 0.2 | 1.2 | 0.0 | 0.0 | 0.1 |
| 73 | 3.1 | 0.2 | 1.2 | 0.2 | 0.0 | 0.2 |
| 74 | 1.2 | 0.0 | 0.7 | 0.0 | 0.0 | 0.1 |
| 75 | 0.2 | 0.0 | 0.6 | 0.0 | 0.0 | 0.1 |
| 76 | 0.1 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 |
| 77 | 0.6 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 |
| 78 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 |
| 79 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 |
| 80 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |

Table 7.4.3. Overview of the number of OTB (otter trawl) discard trips, the total number of OTB trips and the raising factor used to raise the Irish discard data to international discards.




[^13]Table 7.4.4. (a) Catch numbers-at-age of haddock in VIIb-k. (b) Landings numbers-at-age. (c) Discard numbers-at-age. Strong year classes are highlighted.

| a) Haddock VIlbk - Landings numbers at age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 9 | 494 | 3311 | 954 | 815 | 257 | 130 | 130 | 42 | 3 | 0 \#1993 |
| 0 | 1491 | 2934 | 870 | 461 | 297 | 66 | 25 | 63 | 0 | 0 \#1994 |
| 25 | 2237 | 1185 | 1090 | 462 | 581 | 338 | 161 | 44 | 0 | 0 \#1995 |
| 0 | 2399 | 10373 | 1206 | 648 | 260 | 275 | 126 | 71 | 10 | 10 \#1996 |
| 0 | 1581 | 12102 | 3119 | 694 | 580 | 239 | 130 | 33 | 42 | 22 \#1997 |
| 3 | 640 | 3264 | 6199 | 846 | 302 | 252 | 179 | 73 | 56 | 6 \#1998 |
| 0 | 622 | 2585 | 1560 | 1646 | 245 | 80 | 44 | 14 | 3 | 0 \#1999 |
| 28 | 4676 | 2344 | 587 | 535 | 589 | 134 | 23 | 14 | 2 | 0 \#2000 |
| 11 | 3998 | 8036 | 1053 | 282 | 295 | 298 | 51 | 29 | 7 | 0 \#2001 |
| 1 | 872 | 4216 | 3354 | 760 | 39 | 88 | 73 | 19 | 5 | 2 \#2002 |
| 16 | 665 | 8293 | 1998 | 1149 | 112 | 42 | 48 | 41 | 10 | 0 \#2003 |
| 4 | 117 | 5870 | 4540 | 881 | 573 | 50 | 12 | 16 | 3 | 0 \#2004 |
| 0 | 783 | 833 | 4166 | 1884 | 436 | 114 | 4 | 13 | 3 | 0 \#2005 |
| 0 | 831 | 3313 | 1431 | 2106 | 376 | 64 | 7 | 0 | 0 | 0 \#2006 |
| 0 | 653 | 6198 | 2566 | 503 | 827 | 149 | 29 | 3 | 2 | 0 \#2007 |
| 0 | 1528 | 3854 | 4212 | 914 | 216 | 358 | 65 | 11 | 1 | 0 \#2008 |
| 0 | 951 | 8532 | 2934 | 1575 | 437 | 188 | 170 | 24 | 3 | 0 \#2009 |
| b) Haddock VIIbk - Discard numbers at age |  |  |  |  |  |  |  |  |  |  |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 577 | 3092 | 1488 | 95 | 7 | 2 | 0 | 0 | 0 | 0 | 0 \#1993 |
| 577 | 3092 | 1488 | 95 | 7 | 2 | 0 | 0 | 0 | 0 | 0 \#1994 |
| 12740 | 1620 | 81 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \#1995 |
| 192 | 4144 | 1497 | 42 | 19 | 6 | 0 | 0 | 0 | 0 | 0 \#1996 |
| 992 | 5457 | 3167 | 252 | 8 | 1 | 0 | 0 | 0 | 0 | 0 \#1997 |
| 423 | 602 | 534 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \#1998 |
| 607 | 2597 | 460 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \#1999 |
| 4737 | 35484 | 6935 | 291 | 14 | 0 | 0 | 0 | 0 | 0 | 0 \#2000 |
| 1247 | 6913 | 2050 | 199 | 14 | 1 | 0 | 0 | 0 | 0 | 0 \#2001 |
| 11949 | 22165 | 6810 | 978 | 60 | 4 | 0 | 0 | 0 | 0 | 0 \#2002 |
| 11303 | 25087 | 10001 | 395 | 150 | 0 | 0 | 0 | 0 | 0 | 0 \#2003 |
| 1470 | 4365 | 10011 | 1203 | 65 | 79 | 0 | 0 | 0 | 0 | 0 \#2004 |
| 1226 | 3302 | 3136 | 1897 | 78 | 0 | 0 | 0 | 0 | 0 | 0 \#2005 |
| 6091 | 5108 | 656 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \#2006 |
| 2171 | 6532 | 4052 | 306 | 5 | 5 | 0 | 0 | 0 | 0 | 0 \#2007 |
| 2658 | 29246 | 8653 | 1016 | 40 | 0 | 0 | 0 | 0 | 0 | 0 \#2008 |
| 6980 | 16502 | 7447 | 329 | 41 | 0 | 0 | 0 | 0 | 0 | 0 \#2009 |
| c) Haddock VIIbk - Catch numbers at age |  |  |  |  |  |  |  |  |  |  |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 585 | 3586 | 4799 | 1049 | 822 | 259 | 130 | 130 | 42 | 3 | 0 \#1993 |
| 577 | 4583 | 4422 | 965 | 468 | 299 | 66 | 25 | 63 | 0 | 0 \#1994 |
| 12766 | 3857 | 1265 | 1090 | 462 | 581 | 338 | 161 | 44 | 0 | 0 \#1995 |
| 192 | 6543 | 11870 | 1248 | 667 | 266 | 275 | 126 | 71 | 10 | 10 \#1996 |
| 992 | 7038 | 15269 | 3372 | 702 | 581 | 239 | 130 | 33 | 42 | 22 \#1997 |
| 425 | 1242 | 3798 | 6232 | 846 | 302 | 252 | 179 | 73 | 56 | 6 \#1998 |
| 607 | 3218 | 3045 | 1568 | 1646 | 245 | 80 | 44 | 14 | 3 | 0 \#1999 |
| 4765 | 40160 | 9279 | 879 | 549 | 589 | 134 | 23 | 14 | 2 | 0 \#2000 |
| 1257 | 10911 | 10086 | 1252 | 296 | 296 | 298 | 51 | 29 | 7 | 0 \#2001 |
| 11950 | 23037 | 11026 | 4331 | 820 | 43 | 88 | 73 | 19 | 5 | 2 \#2002 |
| 11319 | 25752 | 18294 | 2392 | 1299 | 112 | 42 | 48 | 41 | 10 | 0 \#2003 |
| 1474 | 4482 | 15881 | 5742 | 947 | 652 | 50 | 12 | 16 | 3 | 0 \#2004 |
| 1226 | 4085 | 3969 | 6062 | 1962 | 436 | 114 | 4 | 13 | 3 | 0 \#2005 |
| 6091 | 5939 | 3969 | 1431 | 2106 | 376 | 64 | 7 | 0 | 0 | 0 \#2006 |
| 2171 | 7186 | 10250 | 2871 | 508 | 832 | 149 | 29 | 3 | 2 | 0 \#2007 |
| 2658 | 30774 | 12507 | 5229 | 954 | 216 | 358 | 65 | 11 | 1 | 0 \#2008 |
| 6980 | 17453 | 15979 | 3263 | 1616 | 437 | 188 | 170 | 24 | 3 | 0 \#2009 |

Table 7.4.5. (a) Mean landings weights-at-age. (b) Mean discard weights-at-age. (c) Mean stock weights-at-age (including discards). A 3-year running average was applied to the stock weights.

| a) Haddock VIIbk - Landings weights at age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 0.141 | 0.187 | 0.320 | 0.556 | 0.851 | 1.402 | 1.693 | 2.130 | 2.593 | 2.325 | 2.325 \#1993 |
| 0.000 | 0.321 | 0.537 | 0.869 | 1.167 | 1.428 | 1.990 | 2.399 | 2.673 | 2.593 | 2.325 \#1994 |
| 0.156 | 0.285 | 0.735 | 0.932 | 0.964 | 1.052 | 1.284 | 2.040 | 2.495 | 2.673 | 2.593 \#1995 |
| 0.000 | 0.207 | 0.339 | 0.689 | 1.137 | 1.389 | 1.450 | 1.850 | 2.105 | 1.835 | 1.415 \#1996 |
| 0.000 | 0.321 | 0.442 | 0.863 | 1.237 | 1.417 | 1.453 | 0.965 | 1.451 | 0.706 | 1.570 \#1997 |
| 0.101 | 0.291 | 0.341 | 0.664 | 1.024 | 1.325 | 1.558 | 1.915 | 2.106 | 1.544 | 2.044 \#1998 |
| 0.000 | 0.360 | 0.444 | 0.661 | 1.094 | 1.406 | 2.267 | 2.594 | 2.559 | 1.575 | 1.544 \#1999 |
| 0.160 | 0.437 | 0.918 | 1.392 | 1.709 | 1.826 | 2.308 | 2.486 | 2.213 | 2.449 | 1.575 \#2000 |
| 0.442 | 0.345 | 0.541 | 1.104 | 1.865 | 1.783 | 1.705 | 2.297 | 1.669 | 1.386 | 2.449 \#2001 |
| 0.114 | 0.373 | 0.513 | 0.825 | 1.032 | 1.732 | 1.671 | 1.504 | 1.532 | 1.589 | 1.840 \#2002 |
| 0.282 | 0.347 | 0.520 | 0.883 | 1.242 | 1.429 | 1.800 | 1.705 | 1.589 | 2.143 | 3.045 \#2003 |
| 0.197 | 0.432 | 0.523 | 0.758 | 1.192 | 1.380 | 1.855 | 1.806 | 1.876 | 3.092 | 1.950 \#2004 |
| 0.104 | 0.429 | 0.546 | 0.719 | 1.027 | 1.256 | 1.946 | 2.667 | 1.881 | 2.185 | 2.708 \#2005 |
| 0.000 | 0.349 | 0.482 | 0.545 | 0.938 | 1.486 | 2.118 | 2.619 | 4.022 | 4.019 | 2.185 \#2006 |
| 0.000 | 0.330 | 0.467 | 0.640 | 0.886 | 1.199 | 1.630 | 1.487 | 3.427 | 1.448 | 5.779 \#2007 |
| 0.000 | 0.377 | 0.519 | 0.673 | 0.875 | 1.139 | 1.267 | 1.654 | 1.745 | 2.553 | 2.878 \#2008 |
| 0.000 | 0.360 | 0.541 | 0.796 | 1.037 | 1.185 | 1.402 | 1.479 | 2.040 | 1.307 | 0.000 \#2009 |
| b) Haddock VIIbk - Discard weights at age |  |  |  |  |  |  |  |  |  |  |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 0.074 | 0.184 | 0.384 | 0.538 | 0.305 | 0.329 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1993 |
| 0.074 | 0.184 | 0.384 | 0.538 | 0.305 | 0.329 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1994 |
| 0.095 | 0.283 | 0.166 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1995 |
| 0.056 | 0.166 | 0.454 | 0.471 | 0.551 | 0.830 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1996 |
| 0.082 | 0.144 | 0.369 | 0.639 | 0.255 | 0.490 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1997 |
| 0.069 | 0.244 | 0.360 | 0.505 | 0.255 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1998 |
| 0.059 | 0.176 | 0.357 | 0.551 | 0.163 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#1999 |
| 0.091 | 0.134 | 0.325 | 0.200 | 0.198 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2000 |
| 0.096 | 0.166 | 0.347 | 0.435 | 0.553 | 0.322 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2001 |
| 0.084 | 0.211 | 0.341 | 0.407 | 1.333 | 1.174 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2002 |
| 0.012 | 0.192 | 0.318 | 0.246 | 0.750 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2003 |
| 0.085 | 0.207 | 0.366 | 0.586 | 0.907 | 0.523 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2004 |
| 0.068 | 0.187 | 0.317 | 0.473 | 0.694 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2005 |
| 0.066 | 0.376 | 0.259 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2006 |
| 0.088 | 0.201 | 0.433 | 0.593 | 0.335 | 0.206 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2007 |
| 0.077 | 0.174 | 0.416 | 0.583 | 0.224 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2008 |
| 0.089 | 0.233 | 0.410 | 0.550 | 0.311 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 \#2009 |
| c) Haddock VIlbk - Stock weights at age (3-year running average) |  |  |  |  |  |  |  |  |  |  |
| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 0.073 | 0.190 | 0.362 | 0.659 | 0.866 | 1.581 | 2.010 | 2.114 | 3.779 | 2.384 | 2.384 \#1993 |
| 0.080 | 0.252 | 0.441 | 0.766 | 0.956 | 1.536 | 1.864 | 2.142 | 3.403 | 2.473 | 2.473 \#1994 |
| 0.075 | 0.244 | 0.438 | 0.821 | 1.147 | 1.385 | 1.720 | 1.998 | 2.833 | 2.166 | 2.170 \#1995 |
| 0.078 | 0.238 | 0.454 | 0.936 | 1.266 | 1.484 | 1.712 | 1.961 | 2.484 | 1.575 | 2.033 \#1996 |
| 0.069 | 0.173 | 0.342 | 0.805 | 1.234 | 1.421 | 1.676 | 2.044 | 2.542 | 1.037 | 1.368 \#1997 |
| 0.070 | 0.180 | 0.366 | 0.671 | 1.071 | 1.311 | 1.782 | 2.198 | 2.231 | 1.040 | 1.365 \#1998 |
| 0.073 | 0.182 | 0.363 | 0.645 | 1.088 | 1.456 | 2.050 | 2.363 | 2.324 | 1.376 | 1.358 \#1999 |
| 0.082 | 0.185 | 0.430 | 0.785 | 1.335 | 1.691 | 2.224 | 2.373 | 1.934 | 1.560 | 1.560 \#2000 |
| 0.090 | 0.195 | 0.410 | 0.864 | 1.380 | 1.710 | 1.893 | 1.817 | 1.562 | 1.753 | 1.670 \#2001 |
| 0.064 | 0.199 | 0.399 | 0.827 | 1.372 | 1.671 | 1.895 | 1.724 | 1.601 | 2.647 | 2.409 \#2002 |
| 0.060 | 0.204 | 0.367 | 0.727 | 1.200 | 1.556 | 1.944 | 1.386 | 1.610 | 3.145 | 2.409 \#2003 |
| 0.055 | 0.207 | 0.352 | 0.682 | 1.225 | 1.626 | 2.315 | 1.968 | 1.927 | 3.147 | 2.732 \#2004 |
| 0.073 | 0.260 | 0.358 | 0.604 | 1.100 | 1.555 | 2.172 | 2.421 | 2.676 | 3.151 | 2.891 \#2005 |
| 0.074 | 0.261 | 0.369 | 0.580 | 0.982 | 1.448 | 2.110 | 2.564 | 3.343 | 2.752 | 4.182 \#2006 |
| 0.077 | 0.247 | 0.391 | 0.590 | 0.885 | 1.361 | 1.820 | 2.203 | 3.315 | 3.079 | 4.179 \#2007 |
| 0.085 | 0.211 | 0.406 | 0.632 | 0.897 | 1.196 | 1.641 | 1.847 | 3.304 | 2.287 | 3.387 \#2008 |
| 0.083 | 0.212 | 0.406 | 0.630 | 0.891 | 1.147 | 1.488 | 1.825 | 3.116 | 2.319 | 2.182 \#2009 |

Table 7.4.6. Tuning data available for haddock in VIIB-k. The tuning data used in the final assessment is highlighted in grey.

HADDOCK VIIb-k, WGSSDS 2009, TUNING DATA, updated HG 280410

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IR-7b-OT : Irish Otter Trawl in 7B - effort, nos at age per 1000h
19952009

| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 10 |  |  |  |  |  |  |  |  |  |
| 65.3 | 0 | 0 | 20.5 | 104.3 | 76.1 | 105.3 | 62 | 29.6 | 8.1 | 0 |
|  | 0 | \#1995 | 423.4 t |  |  |  |  |  |  |  |
| 41.5 | 0 | 19.4 | 93.2 | 30.2 | 30 | 17.9 | 21.5 | 9.4 | 5.1 | 0.8 |
|  | 0.8 | \#1996 | 187.0 t |  |  |  |  |  |  |  |
| 49.5 | 0 | 8.3 | 195.2 | 116.9 | 29.6 | 31.9 | 19.1 | 13.5 | 4.1 | 5.3 |
|  | 8.4 | \#1997 | 273.3 t |  |  |  |  |  |  |  |
| 63.5 | $\bigcirc$ | 9.8 | 147.4 | 290.7 | 68.1 | 37.7 | 34.6 | 25 | 9.5 | 8.4 |
|  | 0.9 | \#1998 | 445.2 t |  |  |  |  |  |  |  |
| 62 | 0 | 0.4 | 193.6 | 225.9 | 190.9 | 49.6 | 12.4 | 6 | 2.3 | 0.7 |
|  | 0 | \#1999 | 404.2 t |  |  |  |  |  |  |  |
| 57.7 | 0 | 41.3 | 57.2 | 22.2 | 56.8 | 98.5 | 31.2 | 7.5 | 6.9 | 0.7 |
|  | 0 | \#2000 | 299.8 t |  |  |  |  |  |  |  |
| 60.7 | 0.0 | 20.2 | 289.1 | 72.8 | 13.9 | 42.5 | 60.4 | 7.4 | 8.2 | 2.0 |
|  | 0.0 | \#2001 | 298.9 t |  |  |  |  |  |  |  |
| 46.8 | 0.26 | 3.9 | 38.9 | 95.2 | 28.6 | 4.3 | 17.3 | 17.6 | 4.8 | 1.3 |
|  | 0.6 | \#2002 | 160.0 t |  |  |  |  |  |  |  |
| 64.0 | 0.0 | 2.2 | 21.7 | 42.2 | 66.8 | 15.1 | 9.0 | 10.6 | 10.4 | 2.5 |
|  | 0.1 | \#2003 | 163.6 t |  |  |  |  |  |  |  |
| 60.4 | 0.0 | 0.6 | 43.7 | 68.3 | 59.8 | 79.6 | 11.0 | 3.2 | 4.8 | 0.3 |
|  | 0.2 | \#2004 | 189.0 t |  |  |  |  |  |  |  |
| 47.4 | 0.0 | 9.7 | 60.8 | 64.4 | 57.4 | 32.7 | 2.0 | 1.6 | 1.0 | 0.3 |
|  | 0.0 | \#2005 | 157.5 t |  |  |  |  |  |  |  |
| 39.7 | 0.0 | 20.9 | 120.5 | 108.9 | 50.7 | 7.2 | 9.3 | 0.0 | 0.0 | 0.0 |
|  | 0.0 | \#2006 | 141.9 t |  |  |  |  |  |  |  |
| 40.7 | 0.0 | 0.0 | 63.5 | 64.9 | 45.3 | 69.5 | 14.9 | 7.9 | 0.0 | 0.0 |
|  | 0.0 | \#2007 | 192.6 t |  |  |  |  |  |  |  |
| 37.3 | 0.0 | 0.0 | 37.6 | 96.6 | 63.3 | 33.3 | 62.4 | 12.2 | 3.1 | 0.2 |
|  | 0.0 | \#2008 | 202.9 t |  |  |  |  |  |  |  |
| 37.8 | 0.0 | 0.1 | 75.1 | 54.3 | 81.3 | 80.5 | 34.4 | 44.3 | 4.7 | 1.3 |
|  | 0.0 | \#2009 | 253.7 t |  |  |  |  |  |  |  |

IR-7j-OT : Irish Otter Trawl in 7J - effort, nos at age per 1000h
19952009

| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 10 |  |  |  |  |  |  |  |  |  |
| 93.6 | 3.56 | 323.2 | 92.2 | 37.7 | 1.4 | 0.5 | $\bigcirc$ | 0 | $\bigcirc$ | $\bigcirc$ |
|  | 0 | \#1995 | 220.7 t |  |  |  |  |  |  |  |


| 70.2 | 0 | 146.9 | 464.1 | 24 | 9.9 | 3.2 | 1.6 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0 | $\# 1996$ | 236.0 t |  |  |  |  |  |  |  |
| 83.2 | 0 | 136.4 | 929 | 190.9 | 38.6 | 26.4 | 6.7 | 1.5 | 0 | 0 |
|  | 0 | $\# 1997$ | 758.5 t |  |  |  |  |  |  |  |
| 89.6 | 0.34 | 69 | 287.7 | 515.6 | 48 | 7.3 | 4.3 | 3 | 1.6 | 0 |
|  | 0 | $\# 1998$ | 581.9 t |  |  |  |  |  |  |  |
| 40.6 | 0 | 8.5 | 119.2 | 52.1 | 61.2 | 3.2 | 1.6 | 1.8 | 0.6 | 0 |
|  | 0 | $\# 1999$ | 183.9 t |  |  |  |  |  |  |  |
| 64.1 | 0 | 100.1 | 80.4 | 30.6 | 26.2 | 37 | 4.9 | 0 | 0 | 0 |
|  | 0 | $\# 2000$ | 305.3 t |  |  |  |  |  |  |  |
| 67.7 | 0.4 | 347.9 | 523.0 | 62.7 | 21.1 | 10.4 | 6.3 | 1.4 | 0.1 | 0.0 |
|  | 0.0 | $\# 2001$ | 564.0 t |  |  |  |  |  |  |  |
| 90.4 | 0.2 | 38.9 | 495.4 | 322.3 | 36.0 | 3.9 | 7.3 | 3.2 | 0.6 | 0.0 |
|  | 0.0 | $\# 2002$ | 587.3 t |  |  |  |  |  |  |  |
| 111.3 | 0.7 | 26.6 | 318.3 | 125.7 | 150.1 | 23.0 | 3.6 | 4.1 | 2.6 | 0.0 |
|  | 0.0 | $\# 2003$ | 483.4 t |  |  |  |  |  |  |  |
| 92.0 | 0.0 | 7.8 | 204.5 | 207.1 | 84.4 | 34.4 | 2.4 | 0.8 | 0.6 | 0.3 |
|  | 0.0 | $\# 2004$ | 362.5 t |  |  |  |  |  |  |  |
| 73.9 | 0.1 | 2.3 | 32.2 | 207.1 | 152.6 | 61.2 | 9.6 | 0.0 | 0.0 | 0.0 |
|  | 0.0 | $\# 2005$ | 339.5 t |  |  |  |  |  |  |  |
| 65.9 | 0.0 | 32.4 | 117.6 | 111.7 | 222.8 | 44.3 | 5.4 | 0.9 | 0.0 | 0.0 |
|  | 0.0 | $\# 2006$ | 333.9 t |  |  |  |  |  |  |  |
| 80.5 | 0.0 | 28.1 | 148.6 | 152.6 | 41.9 | 157.8 | 16.6 | 2.1 | 0.6 | 0.0 |
|  | 0.2 | $\# 2007$ | 386.7 t |  |  |  |  |  |  |  |
| 66.5 | 0.0 | 177.7 | 232.8 | 120.6 | 74.4 | 22.6 | 38.5 | 8.3 | 0.5 | 0.0 |
|  | 0.1 | $\# 2008$ | 379.3 t |  |  |  |  |  |  |  |
| 72.5 | 0.0 | 102.0 | 577.5 | 105.6 | 52.5 | 38.6 | 34.8 | 20.4 | 3.1 | 0.0 |
|  | 0.0 | $\# 2009$ | 572.0 t |  |  |  |  |  |  |  |

IR-7bj-OT : Irish Otter Trawl in 7B\&J - effort, nos at age per 1000h $1995 \quad 2009$
$\begin{array}{llll}1 & 1 & 0 & 1\end{array}$
$0 \quad 10$

| 158.9 | 3.56 | 323.2 | 112.7 | 142 | 77.6 | 105.8 | 62 | 29.6 | 8.1 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | \#1995 | 644.1 t |  |  |  |  |  |  |  |
| 111.7 | 0 | 166.3 | 557.4 | 54.1 | 39.9 | 21.1 | 23.1 | 9.4 | 5.1 | 0.8 |
|  | 0.8 | \#1996 | 423.0 t |  |  |  |  |  |  |  |
| 132.7 | 0 | 144.7 | 1124.2 | 307.8 | 68.1 | 58.2 | 25.8 | 15 | 4.1 | 5.3 |
|  | 8.4 | \#1997 | 1031.8 |  |  |  |  |  |  |  |
| 153.1 | 0.34 | 78.8 | 435.1 | 806.3 | 116.1 | 45.1 | 39 | 28 | 11.2 | 8.4 |
|  | 0.9 | \#1998 | 1027.1 |  |  |  |  |  |  |  |
| 102.7 | 0 | 8.9 | 312.8 | 277.9 | 252.1 | 52.8 | 13.9 | 7.8 | 3 | 0.7 |
|  | 0 | \#1999 | 588.1 t |  |  |  |  |  |  |  |
| 121.7 | 0 | 141.3 | 137.6 | 52.8 | 83 | 135.5 | 36.1 | 7.5 | 6.9 | 0.7 |
|  | 0 | \#2000 | 605.1 t |  |  |  |  |  |  |  |
| 128.4 | 0.4 | 368.1 | 812.0 | 135.6 | 35.0 | 52.9 | 66.7 | 8.8 | 8.3 | 2.0 |
|  | 0.0 | \#2001 | 862.9 t |  |  |  |  |  |  |  |
| 137.2 | 0.5 | 42.9 | 534.2 | 417.5 | 64.6 | 8.3 | 24.6 | 20.8 | 5.4 | 1.3 |
|  | 0.6 | \#2002 | 747.3 t |  |  |  |  |  |  |  |
| 175.2 | 0.7 | 28.8 | 340.0 | 167.9 | 216.9 | 38.1 | 12.6 | 14.7 | 13.0 | 2.5 |
|  | 0.1 | \#2003 | 647.0 |  |  |  |  |  |  |  |


| 152.4 | 0.0 | 8.4 | 248.2 | 275.3 | 144.2 | 114.0 | 13.4 | 4.0 | 5.4 | 0.6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.2 | \#2004 | 551.5 t |  |  |  |  |  |  |  |
| 121.3 | 0.1 | 12.1 | 92.9 | 271.6 | 210.1 | 93.9 | 11.7 | 1.6 | 1.0 | 0.3 |
|  | 0.0 | \#2005 | 497.0 t |  |  |  |  |  |  |  |
| 105.6 | 0.0 | 53.3 | 238.1 | 220.6 | 273.6 | 51.5 | 14.7 | 0.9 | 0.0 | 0.0 |
|  | 0.0 | \#2006 | 475.8 t |  |  |  |  |  |  |  |
| 121.2 | 0.0 | 28.1 | 212.0 | 217.5 | 87.2 | 227.3 | 31.5 | 10.0 | 0.6 | 0.0 |
|  | 0.2 | \#2007 | 579.3 t |  |  |  |  |  |  |  |
| 103.8 | 0.0 | 177.7 | 270.4 | 217.2 | 137.7 | 56.0 | 100.9 | 20.5 | 3.6 | 0.2 |
|  | 0.1 | \#2008 | 582.2 t |  |  |  |  |  |  |  |
| 110.3 | 0.0 | 102.1 | 652.6 | 160.0 | 133.7 | 119.1 | 69.2 | 64.7 | 7.8 | 1.3 |
|  | 0.0 | \#2009 | 825.7 t |  |  |  |  |  |  |  |

IR-7g-ISCSGFS : Irish Sea Celtic Sea GFS (VIIg; Prime stations only) - effort, nos at age per 30min 19972002

| 1 | 1 | 0.8 | 0.9 |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 4 |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 1 | 18.9 | 11.7 | 15.2 | 2.4 | 2.4 | \#1997 |
|  |  |  |  |  |  |  |
| 1 | 241.6 | 23.6 | 5.6 | 0.8 | 0.2 | \#1998 |
| 1 | 2465.2 | 6.6 | 0.4 | 0.4 | 0.1 | \#1999 |
| 1 | 1191.4 | 710.6 | 0.9 | 0 | 0 | \#2000 |
| 1 | 1200.9 | 34.5 | 13.7 | 0 | 0 | \#2001 |
| 1 | 560.9 | 119.9 | 8.5 | 2.8 | 0.2 | \#2002 |

IR-7bj-WCGFS : Irish Autum WCGFS - effort, nos at age per min 19932002

| 1 | 1 | 0.75 | 0.79 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 6 |  |  |  |  |  |  |  |
| 1901 | 6647 | 1307 | 86 | 52 | 7 | 6 | 0 | \#1993 |
| 2386 | 47261 | 727 | 111 | 68 | 5 | 7 | 0 | \#1994 |
| 2210 | 239176 | 6136 | 17 | 6 | 2 | 3 | 0 | \#1995 |
| 2248 | 37211 | 9305 | 333 | 141 | 28 | 22 | 0 | \#1996 |
| 2396 | 661 | 8679 | 526 | 249 | 88 | 120 | 0 | \#1997 |
| 2486 | 12340 | 601 | 685 | 451 | 50 | 31 | 0 | \#1998 |
| 2304 | 53123 | 808 | 22 | 66 | 7 | 18 | 0 | \#1999 |
| 2400 | 57484 | 14036 | 28 | 22 | 6 | 22 | 0 | \#2000 |
| 1107 | 45261 | 10419 | 6230 | 209 | 173 | 364 | 302 | \#2001 |
| 1301 | 141437 | 17366 | 2026 | 849 | 7 | 5 | 27 | \#2002 |

UK-7efghj-WCGFS-1gp : Standardised no $<=26 \mathrm{~cm}$ as proxy for 1-gp
19922001

| 1 | 1 | 0.15 | 0.25 |
| :--- | :--- | :--- | :--- |
| 1 | 1 |  |  |
| 1 | 1.7 | $\# 1992$ |  |
| 1 | 19.8 | $\# 1993$ |  |
| 1 | 33.4 | $\# 1994$ |  |
| 1 | 20.8 | $\# 1995$ |  |
| 1 | 145.9 | $\# 1996$ |  |
| 1 | 26.7 | $\# 1997$ |  |
| 1 | 7.1 | $\# 1998$ |  |
| 1 | 9.3 | $\# 1999$ |  |
| 1 | 19.6 | $\# 2000$ |  |



IR-7g-SAGFS : VIIg, Irish Sea Celtic Sea GFS +Irish Groundfish Survey (IBTS 4th Qtr) - effort, nos at age per 10 km 2
19992009
$\begin{array}{llll}1 & 1 & 0.8 & 0.9\end{array}$
$0 \quad 8$


IR-7g-GFS : Irish Groundfish Survey in VIIg (IBTS 4th Qtr) - Haddock no. @ age 20032009

| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 7 |  |  |  |  |  |  |  |  |

IR-7j-GFS : Irish Groundfish Survey in VIIj (IBTS 4th Qtr) - Haddock no. @ age
20032009

| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 7 |  |  |  |  |  |  |  |  |
| 780 | 4592 | 16281 | 640 | 74 | 20 | 1 | 0 | 0 | \#2003 |
| 720 | 5175 | 1620 | 1395 | 44 | 7 | 4 | 1 | 0 | \#2004 |
| 881 | 1474 | 1273 | 240 | 286 | 36 | 6 | 2 | 0 | \#2005 |
| 901 | 2636 | 262 | 124 | 53 | 50 | 7 | 0 | 0 | \#2006 |
| 874 | 22831 | 2116 | 192 | 71 | 20 | 36 | 1 | 0 | \#2007 |
| 873 | 14056 | 4934 | 222 | 20 | 15 | 6 | 6 | 3 | \#2008 |
| 747 | 56856 | 1476 | 205 | 2 | 1 | 2 | 2 | 1 | \#2009 |

IR-7gj-GFS : Irish Groundfish Survey in VIIg \& j (IBTS 4th Qtr) - Haddock no. @ age 20032009

| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 7 |  |  |  |  |  |  |  |  |
| 1612 | 7708 | 23095 | 1212 | 88 | 27 | 1 | 1 | 0 | \#2003 |
| 1740 | 19162 | 4533 | 3109 | 183 | 15 | 10 | 1 | 0 | \#2004 |
| 1726 | 16119 | 5196 | 433 | 413 | 56 | 6 | 2 | 0 | \#2005 |
| 1947 | 7776 | 3433 | 416 | 87 | 75 | 10 | 1 | 0 | \#2006 |
| 2042 | 38414 | 3527 | 611 | 171 | 26 | 38 | 1 | 0 | \#2007 |
| 2012 | 26721 | 9403 | 376 | 62 | 25 | 5 | 11 | 3 | \#2008 |
| 1765 | 145133 | 3014 | 931 | 47 | 8 | 3 | 2 | 1 | \#2009 |

IR-7b-GFS : Irish Groundfish Survey in VIIb (IBTS 4th Qtr) - Haddock no. @ age 20032009

| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 8 |  |  |  |  |  |  |  |  |  |
| 757 | 11834 | 34773 | 2793 | 874 | 313 | 6 | 1 | 2 | 7 | \#2003 |
| 728 | 31311 | 2960 | 6688 | 925 | 372 | 196 | 46 | 2 | 1 | \#2004 |
| 724 | 3737 | 7082 | 964 | 2299 | 188 | 37 | 5 | 0 | 0 | \#2005 |
| 700 | 8823 | 2303 | 2471 | 614 | 421 | 39 | 16 | 7 | 0 | \#2006 |
| 734 | 56350 | 2383 | 770 | 747 | 434 | 392 | 26 | 9 | 0 | \#2007 |
| 653 | 10948 | 11622 | 398 | 148 | 172 | 98 | 273 | 54 | 4 | \#2008 |
| 770 | 46145 | 6349 | 8264 | 258 | 272 | 122 | 165 | 110 | 4 | \#2009 |

Table 7.4.7. Lpue of haddock and effort for Irish Otter trawls in VIIb, VIIg and VIIj, the French gadoid fleet in VIIfgh and effort only for UK beam and trawl fleets in VIIe-k. Lpue in kg/hour and effort in hours fishing.

|  | $\begin{gathered} \hline \text { IRL OTB } \\ \text { VIIb } \\ \hline \end{gathered}$ |  | $\begin{gathered} \hline \text { IRL OTB } \\ \text { VIIg } \\ \hline \end{gathered}$ |  | $\begin{gathered} \hline \text { IRL OTB } \\ \text { VIIj } \\ \hline \end{gathered}$ |  | $\begin{gathered} \text { FR GAD } \\ \text { VIIfgh } \\ \hline \end{gathered}$ |  | JK BeanJK Traw VIle-k VIle-k |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | LPUE | Effort | LPUE | Effort | LPUE | Effort | LPUE | Effort | Effort | Effort |
| 1983 |  |  |  |  |  |  | 2.18 | 115379 | 135344 | 82054 |
| 1984 |  |  |  |  |  |  | 2.02 | 85790 | 131465 | 86722 |
| 1985 |  |  |  |  |  |  | 2.83 | 92012 | 152487 | 90298 |
| 1986 |  |  |  |  |  |  | 1.64 | 119664 | 135738 | 84748 |
| 1987 |  |  |  |  |  |  | 3.20 | 144186 | 177118 | 84267 |
| 1988 |  |  |  |  |  |  | 7.27 | 221164 | 194882 | 89148 |
| 1989 |  |  |  |  |  |  | 5.28 | 247929 | 198156 | 84140 |
| 1990 |  |  |  |  |  |  | 2.23 | 201349 | 207576 | 99492 |
| 1991 |  |  |  |  |  |  | 1.94 | 179381 | 203196 | 76712 |
| 1992 |  |  |  |  |  |  | 3.74 | 190784 | 196065 | 86397 |
| 1993 |  |  |  |  |  |  | 4.23 | 213508 | 208421 | 61903 |
| 1994 |  |  |  |  |  |  | 7.95 | 181031 | 220023 | 53743 |
| 1995 | 6.47 | 65423 | 1.48 | 63560 | 2.36 | 93688 | 9.12 | 184067 | 243136 | 52270 |
| 1996 | 4.51 | 41496 | 5.36 | 60041 | 3.36 | 70237 | 15.36 | 170141 | 260817 | 60509 |
| 1997 | 5.51 | 49560 | 5.82 | 65105 | 9.12 | 83187 | 19.58 | 226015 | 264814 | 66707 |
| 1998 | 7.00 | 63560 | 4.09 | 72298 | 6.49 | 89610 | 11.62 | 189457 | 254590 | 62114 |
| 1999 | 6.51 | 62047 | 2.34 | 51657 | 4.53 | 40609 | 5.05 | 206601 | 251431 | 98350 |
| 2000 | 5.05 | 62758 | 10.43 | 60604 | 4.68 | 64626 | 8.86 | 170292 | 258962 | 104088 |
| 2001 | 4.92 | 60725 | 8.34 | 69427 | 8.34 | 67659 | 16.39 | 190482 | 272662 | 85338 |
| 2002 | 3.42 | 46793 | 3.28 | 77689 | 6.49 | 90446 | 13.61 | 176678 | 249480 | 83023 |
| 2003 | 2.56 | 63959 | 3.28 | 86791 | 4.34 | 111267 | 22.01 | 144180 | 282097 | 72303 |
| 2004 | 3.13 | 60446 | 3.45 | 96991 | 3.94 | 91957 | 31.41 | 119444 | 273871 | 75681 |
| 2005 | 3.32 | 47399 | 4.42 | 124395 | 4.59 | 73920 | 21.48 | 101027 | 270347 | 76361 |
| 2006 | 3.58 | 39698 | 4.16 | 119227 | 5.07 | 65856 | 17.74 | 79214 | 252001 | 83308 |
| 2007 | 4.73 | 40718 | 4.10 | 136525 | 4.80 | 80485 | 22.62 | 83904 | 239921 | 87683 |
| 2008 | 5.44 | 37338 | 4.57 | 125815 | 5.70 | 66503 | 31.22 | 70044 | 216529 | 71154 |
| 2009 | 6.71 | 37805 | 9.51 | 135178 | 7.89 | 72453 |  |  | 190914 | 73847 |

LPUE in kg/hour fishing
Effort in hours fishing

Table 7.4.8. XSA diagnostics for haddock in VIIb-k.

```
Lowestoft VPA Version 3.1
    30/04/2010 14:56
Extended Survivors Analysis
HADDOCK VIIb-k, WGCSE 2010, COMBSEX, PLUSGROUP
CPUE data from file had7bktu.txt
Catch data for 17 years. 1993 to 2009. Ages 0 to 8.
    Fleet, First, Last, First, Last, Alpha, Beta
    year, year, age , age
"IR-7bj-0T : Irish 0, 1995, 2009, 2, 7, .000, 1.000
"FR-7fghj-EVHOE: THA, 1997, 2009, 0, 5, .750, 1.000
"FR-7fgh-GAD : Frenc, 2002, 2009, 2, 6, .000, 1.000
"IR-7g-SAGFS : VIIg,, 1999, 2009, 0, 5, .800, . }90
Time series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability independent of stock size for all ages
    Catchability independent of age for ages >= 4
Terminal population estimation :
    Survivor estimates shrunk towards the mean F
    of the final 5 years or the 3 oldest ages.
    S.E. of the mean to which the estimates are shrunk = 1.500
    Minimum standard error for population
    estimates derived from each fleet = . 300
    Prior weighting not applied
    Tuning converged after 30 iterations
1
Regression weights
            , 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000
Fishing mortalities
    Age, 2000, 2001, 2002, 2003, 2004, 2005, 2006, 2007, 2008, 2009
            0, .000, .000, .000, .000, .000, .000, .000, .000, .000, .000
            1, 1.011, .428, .485, .515, .380, . 258, .208, .270, .590, . 560
            2, 1.377, .769, 1.078, .929, .707, .693, .431, .667, 1.074, . 714
            3, .903, .670, .933, .721, .884, .653, .580, .646, .894, .953
            4, .531, .924, 1.440, . 832, .715, .900, .496, .417, .459, . 788
            5, .545, .618, .314, .770, 1.589, .884, .418, .371, .313, . 394
            6, .532, .595, .372, .581, 1.002, 1.788, .294, .289, .269, . 496
            7, .231, .395, .278, .357, .321, .184, .467, .210, .197, . }19
```

1
XSA population numbers (Thousands)
AGE
$\begin{array}{lllllllll} & \text { YEAR } & \text { 0, } & \text { 1, } & \text { 2, }\end{array}$
$2000,4.23 \mathrm{E}+04,6.98 \mathrm{E}+04,1.37 \mathrm{E}+04,1.63 \mathrm{E}+03,1.47 \mathrm{E}+03,1.55 \mathrm{E}+03,3.59 \mathrm{E}+02,1.23 \mathrm{E}+02$,

```
2001, 8.10E+04, 3.46E+04, 2.08E+04, 2.83E+03, 5.42E+02, 7.10E+02, 7.35E+02, 1.73E+02,
2002 , 8.64E+04, 6.63E+04, 1.85E+04, 7.89E+03, 1.19E+03, 1.76E+02, 3.13E+02, 3.32E+02,
2003 , 1.91E+04, 7.07E+04, 3.34E+04, 5.15E+03, 2.54E+03, 2.30E+02, 1.05E+02, 1.77E+02,
2004 , 2.42E+04, 1.57E+04, 3.46E+04, 1.08E+04, 2.05E+03, 9.05E+02, 8.73E+01, 4.83E+01,
2005 , 4.27E+04, 1.98E+04, 8.77E+03, 1.40E+04, 3.65E+03, 8.21E+02, 1.51E+02, 2.62E+01,
2006, 4.10E+04, 3.50E+04, 1.25E+04, 3.59E+03, 5.95E+03, 1.22E+03, 2.78E+02, 2.07E+01,
2007, 9.32E+04, 3.36E+04, 2.33E+04, 6.67E+03, 1.65E+03, 2.97E+03, 6.56E+02, 1.69E+02,
2008, 5.50E+04, 7.63E+04, 2.10E+04, 9.78E+03, 2.86E+03, 8.88E+02, 1.68E+03, 4.02E+02,
2009, 3.78E+05, 4.50E+04, 3.46E+04, 5.87E+03, 3.28E+03, 1.48E+03, 5.31E+02, 1.05E+03,
```

Estimated population abundance at 1st Jan 2010 $0.00 \mathrm{E}+00,3.10 \mathrm{E}+05,2.10 \mathrm{E}+04,1.39 \mathrm{E}+04,1.85 \mathrm{E}+03,1.22 \mathrm{E}+03,8.18 \mathrm{E}+02,2.65 \mathrm{E}+02$,

Taper weighted geometric mean of the VPA populations:
$4.30 \mathrm{E}+04,2.95 \mathrm{E}+04,1.54 \mathrm{E}+04,5.22 \mathrm{E}+03,2.13 \mathrm{E}+03,8.99 \mathrm{E}+02,3.90 \mathrm{E}+02,1.78 \mathrm{E}+02$,

Standard error of the weighted Log(VPA populations) :

$$
.8666, \quad .6805, \quad .6255, \quad .5783, \quad .5323, \quad .6739, \quad .7638,1.0171,
$$

1

Log catchability residuals.


Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 2, | 3, | 4, | 5, | 6, | 7 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -8.3204, | -7.7169, | -7.4066, | -7.4066, | -7.4066, | -7.4066, |
| S.E(Log q), | .7244, | .5045, | .3859, | .6048, | .5712, | .1117, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 2, | 1.65, | -1.346, | 7.41, | .25, | 15, | 1.17, | -8.32, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3, | 1.18, | -.670, | 7.56, | .51, | 15, | .61, | -7.72, |
| 4, | 1.02, | -.129, | 7.40, | .67, | 15, | .41, | -7.41, |
| 5, | 1.04, | -.205, | 7.16, | .62, | 15, | .58, | -7.14, |
| 6, | 1.38, | -2.082, | 7.49, | .69, | 15, | .55, | -7.06, |
| 7, | .98, | .684, | 7.35, | .99, | 15, | .11, | -7.39, |

1

Fleet : "FR-7fghj-EVHOE: THA


| Age |  | 2000, | 2001, | 2002, | 2003, | 2004, | 2005, | 2006, | 2007, | 2008, | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | -.72, | .61, | .94, | .65, | 1.00, | .72, | -1.26, | -.15, | .18, | -. 19 |
| 1 | 1 | -.37, | -. 08, | -1.00, | 1.88, | -.98, | .76, | -.96, | -.87, | -. 23, | -. 34 |
| 2 | 2 | -1.86, | .10, | 1.37, | 1.08, | . 66, | -. 01, | -. 78, | -. 46, | -.17, | -. 08 |
| 3 | 3 | -1.50, | . 01, | 2.16, | . 43, | . 31 , | .50, | -. 68, | -. 03, | -. 27, | -. 58 |
| 4 | 4 | -1.01, | -1.06, | 2.04, | -. 38, | -.17, | .90, | -.30, | -. 31, | -. 42, | -. 58 |
| 5 | 5 | -1.74, | 99.99, | 99.99, | -. 34, | .11, | .80, | .53, | . 32 , | -.47, | -1.43 |
| 6 | 6 | No dat | for | s fle | at | s age |  |  |  |  |  |
| 7 |  | No data | for th | is flee | t at th | is age |  |  |  |  |  |

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4, | 5 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.4955, | -6.9893, | -8.0627, | -8.4385, | -8.8631, | -8.8631, |
| S.E(Log q), | .7638, | .9370, | .9070, | .9717, | .9143, | 1.0028, |

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope , t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 0, | .91, | .418, | 6.90, | .64, | 13, | .72, | -6.50, |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1, | 1.25, | -.494, | 6.16, | .27, | 13, | 1.21, | -6.99, |
| 2, | 1.03, | -.063, | 8.01, | .30, | 13, | .97, | -8.06, |
| 3, | .57, | 1.833, | 8.54, | .62, | 13, | .51, | -8.44, |
| 4, | .90, | .242, | 8.74, | .36, | 13, | .86, | -8.86, |
| 5, | 1.54, | -.684, | 9.94, | .17, | 10, | 1.59, | -8.86, |

1

Fleet : "FR-7fgh-GAD : Frenc

Age , 2000, 2001, 2002, 2003, 2004, 2005, 2006, 2007, 2008, 2009 0 , No data for this fleet at this age


Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 2, | 3, | 4, | 5, | 6 |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.8843, | -6.2804, | -6.8550, | -6.8550, | -6.8550, |
| S.E(Log q), | .4404, | .4092, | .8520, | .9392, | 1.4644, |

Regression statistics :
Ages with q independent of year class strength and constant w.r.t. time. Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 2, | .63, | 2.005, | 8.00, | .85, | 7, | .23, | -6.88, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3, | .70, | 1.273, | 7.09, | .78, | 7, | .27, | -6.28, |
| 4, | .45, | 2.589, | 7.39, | .82, | 7, | .28, | -6.85, |
| 5, | .68, | 1.421, | 7.00, | .80, | 7, | .54, | -7.20, |
| 6, | 1.17, | -.303, | 8.04, | .38, | 7, | 1.48, | -7.68, |

1


Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4, | 5 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -4.0741, | -4.9157, | -6.0247, | -6.7751, | -7.7020, | -7.7020, |
| S.E(Log q), | .7583, | .8845, | .7386, | .4400, | .6437, | 1.0167, |

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 0, | 1.08, | -.239, | 3.51, | .49, | 11, | .86, | -4.07, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1, | .81, | .446, | 5.97, | .39, | 11, | .75, | -4.92, |
| 2, | .60, | 1.904, | 7.53, | .72, | 11, | .39, | -6.02, |
| 3, | .76, | 1.165, | 7.25, | .74, | 10, | .33, | -6.78, |
| 4, | 2.15, | -1.155, | 7.47, | .13, | 9, | 1.36, | -7.70, |
| 5, | 5.76, | -1.104, | 13.22, | .01, | 6, | 5.04, | -8.15, |

1

Terminal year survivor and $F$ summaries :

Age 0 Catchability constant w.r.t. time and dependent on age
Year class $=2009$

| Fleet, | Estimated, Survivors, | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \end{aligned}$ | Ext, | Var, <br> Ratio, | N, | Scaled, Weights, | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| "IR-7bj-OT : Irish 0, | 1., | .000, | .000, | . 00, | 0, | . 000, | 000 |
| "FR-7fghj-EVHOE: THA, | 255865., | .793, | .000, | . 00, | 1, | 500, | . 000 |
| "FR-7fgh-GAD : Frenc, | 1. | .000, | .000, | . 00, | 0, | . 000, | . 000 |
| "IR-7g-SAGFS : VIIg, | 374972., | .792, | .000, | .00, | 1, | 500, | . 000 |
| F shrinkage mean , | 0., | 1.50, |  |  |  | .000, | . 000 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $309793 .$, | .56, | .19, | 2, | .341, | .000 |

Age 1 Catchability constant w.r.t. time and dependent on age
Year class = 2008

|  | Estimated, | Int, | Ext, | Var, | N, Scaled, Estimated |  |  |
| :---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet, | Survivors, | s.e, | s.e, | Ratio, | , Weights, | F |  |
| "IR-7bj-0T : Irish 0, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "FR-7fghj-EVHOE: THA, | $20464 .$, | .614, | .258, | .42, | 2, | .428, | .572 |
| "FR-7fgh-GAD : Frenc, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "IR-7g-SAGFS : VIIg,, | $18232 .$, | .601, | .305, | .51, | 2, | .447, | .624 |
| F shrinkage mean , | $38548 .$, | $1.50,, 1$, |  |  | .126, | .343 |  |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | S.e, | , | Ratio, |  |
| $21044 .$, | .42, | .18, | 5, | .434, | .560 |

1
Age 2 Catchability constant w.r.t. time and dependent on age Year class = 2007

| Fleet, | Estimated, Survivors, | Int, s.e, | Ext, | Var, <br> Ratio, | N, | Scaled, Weights, | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| "IR-7bj-OT : Irish 0, | 14866., | .748, | . 000, | . 00, | 1, | 210, | . 679 |
| "FR-7fghj-EVHOE: THA, | 12057., | .537, | . 041, | . 08, | 3, | . 305, | . 788 |
| "FR-7fgh-GAD : Frenc, | 1., | . 000, | . 000, | . 00 , | 0, | . 000, | . 000 |
| "IR-7g-SAGFS : VIIg, | 15047., | .495, | .197, | . 40 , | 3, | . 378, | . 673 |
| $F$ shrinkage mean | 13676., | 1.50, |  |  |  | .107, | . 721 |


| $l$ |  |  |  |  |  |
| :--- | ---: | :--- | ---: | ---: | ---: |
| Weighted prediction : |  |  |  |  |  |
| Survivors, | Int, | Ext, | N, | Var, | F |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $13885 .$, | .33, | .08, | 8, | .225, | .714 |

Age 3 Catchability constant w.r.t. time and dependent on age Year class = 2006


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| ---: | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | S.e, | Ratio, |  |  |
| 1854., | .25, | .12, | 12, | .494, | .953 |

1
Age 4 Catchability constant w.r.t. time and dependent on age Year class = 2005


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | ---: | ---: | ---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| 1220., | .20, | .10, | 16, | .520, | .788 |

Age 5 Catchability constant w.r.t. time and age (fixed at the value for age) 4 Year class = 2004



Table 7.4.9. Stock Summary for haddock in VIIb-k.

| Year | Recruits <br> age 0 | TotBio | SSB | Landings | Discards | Yield/ <br> SSB | Fbar 2-5 <br> Lan+Dis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 3}$ | 14212 | 15506 | 11473 | 3348 | 1193 | 0.292 | 0.500 |
| $\mathbf{1 9 9 4}$ | 35576 | 19070 | 13292 | 4131 | 1193 | 0.311 | 0.419 |
| $\mathbf{1 9 9 5}$ | 54814 | 23313 | 12094 | 4470 | 472 | 0.370 | 0.394 |
| $\mathbf{1 9 9 6}$ | 22199 | 31219 | 18807 | 6756 | 1403 | 0.359 | 0.571 |
| $\mathbf{1 9 9 7}$ | 9674 | 24820 | 21008 | 10827 | 2120 | 0.515 | 0.810 |
| $\mathbf{1 9 9 8}$ | 24807 | 18744 | 15582 | 7668 | 356 | 0.492 | 0.775 |
| $\mathbf{1 9 9 9}$ | 85217 | 20190 | 10273 | 4882 | 625 | 0.475 | 0.704 |
| $\mathbf{2 0 0 0}$ | 42285 | 29393 | 13018 | 7411 | 7057 | 0.569 | 0.839 |
| $\mathbf{2 0 0 1}$ | 80956 | 28867 | 14830 | 8632 | 1952 | 0.582 | 0.745 |
| $\mathbf{2 0 0 2}$ | 86396 | 35924 | 17205 | 6403 | 7468 | 0.372 | 0.941 |
| $\mathbf{2 0 0 3}$ | 19138 | 35800 | 20222 | 8146 | 8221 | 0.403 | 0.813 |
| $\mathbf{2 0 0 4}$ | 24213 | 28572 | 23996 | 8581 | 5371 | 0.358 | 0.974 |
| $\mathbf{2 0 0 5}$ | 42737 | 25830 | 17556 | 6555 | 2563 | 0.373 | 0.783 |
| $\mathbf{2 0 0 6}$ | 41018 | 27121 | 14954 | 5383 | 2092 | 0.360 | 0.481 |
| $\mathbf{2 0 0 7}$ | 93185 | 35661 | 20191 | 6510 | 3252 | 0.322 | 0.525 |
| $\mathbf{2 0 0 8}$ | 54954 | 42836 | 22067 | 7049 | 9302 | 0.319 | 0.685 |
| $\mathbf{2 0 0 9}$ | 378381 | 66521 | 25577 | 10028 | 7095 | 0.392 | 0.712 |

Table 7.4.10. Input values for short-term forecast (.prd).


Table 7.4.11. Management options table (.prm).
MFDP version 1a
Run: dp_
Time and date: 10:05 16/05/2010
Fbar age range (Total) : 2-5
Fbar age range Fleet 1 : 2-5

| 2010 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | "CATCH" | Landings | Yield | Discards FBar | Yield |  |  |
| 96800 | 24625 | 1 | 0.4973 | 10860 | 0.1436 | 23383 |  |  |
| 2011 |  |  |  |  |  | 2012 |  |  |
|  |  | "CATCH" | Landings |  | Discards |  |  |  |
| Biomass | SSB | FMult | FBar | Yield | FBar | Yield | Biomass | SSB |
| 87583 | 77906 | 0 | 0 | 0 | 0 | 0 | 115317 | 105640 |
| . | 77906 | 0.1 | 0.0497 | 3293 | 0.0144 | 2864 | 107734 | 98056 |
| . | 77906 | 0.2 | 0.0995 | 6346 | 0.0287 | 5518 | 100718 | 91040 |
| . | 77906 | 0.3 | 0.1492 | 9179 | 0.0431 | 7978 | 94226 | 84549 |
| . | 77906 | 0.4 | 0.1989 | 11806 | 0.0575 | 10259 | 88219 | 78542 |
| . | 77906 | 0.5 | 0.2486 | 14245 | 0.0718 | 12373 | 82660 | 72982 |
| . | 77906 | 0.6 | 0.2984 | 16508 | 0.0862 | 14334 | 77514 | 67836 |
| . | 77906 | 0.7 | 0.3481 | 18610 | 0.1005 | 16153 | 72750 | 63072 |
| . | 77906 | 0.8 | 0.3978 | 20561 | 0.1149 | 17841 | 68338 | 58660 |
| . | 77906 | 0.9 | 0.4476 | 22375 | 0.1293 | 19407 | 64253 | 54575 |
|  | 77906 | 1 | 0.4973 | 24060 | 0.1436 | 20861 | 60468 | 50791 |
| . | 77906 | 1.1 | 0.547 | 25626 | 0.158 | 22211 | 56962 | 47285 |
| . | 77906 | 1.2 | 0.5967 | 27082 | 0.1724 | 23465 | 53714 | 44036 |
|  | 77906 | 1.3 | 0.6465 | 28437 | 0.1867 | 24630 | 50703 | 41025 |
| . | 77906 | 1.4 | 0.6962 | 29698 | 0.2011 | 25712 | 47912 | 38234 |
| . | 77906 | 1.5 | 0.7459 | 30872 | 0.2155 | 26718 | 45324 | 35647 |
|  | 77906 | 1.6 | 0.7956 | 31964 | 0.2298 | 27654 | 42925 | 33247 |
|  | 77906 | 1.7 | 0.8454 | 32982 | 0.2442 | 28524 | 40699 | 31021 |
| . | 77906 | 1.8 | 0.8951 | 33931 | 0.2586 | 29334 | 38634 | 28956 |
|  | 77906 | 1.9 | 0.9448 | 34815 | 0.2729 | 30087 | 36718 | 27040 |
| . | 77906 | 2 | 0.9946 | 35640 | 0.2873 | 30789 | 34939 | 25261 |

Table 7.4.12. Haddock VIIbk. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.

| Y ear-class |  |  | 2007 | 2008 | 2009 | 2010 | 2011 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock No. (thousands) |  |  | 93185 | 54954 | 378381 | 36586 | 36586 |
| of |  | year-olds |  |  |  |  |  |
| Source |  |  | XSA | XSA | XSA | GM | GM |
| Status Quo F: |  |  |  |  |  |  |  |
| \% in | 2010 | landings | 32.9 | 26.0 | 33.5 | 0.0 | - |
| \% in | 2011 |  | 4.3 | 7.8 | 84.7 | 1.8 | 0.0 |
| \% in | 2010 | SSB | 34.8 | 34.3 | 0.0 | 0.0 |  |
| \% in | 2011 | SSB | 5.7 | 6.0 | 81.3 | 0.0 | 0.0 |
| \% in | 2012 | SSB | 5.5 | 4.6 | 67.3 | 14.6 | 0.0 |

Haddock in VIlb-k : Year-class \% contribution to
a) 2011 landings

b) 2012 SSB


Table 7.4.13. Haddock VIIbk. Output from srmsymc ADMB package.

Stock name
had7bk
Sen filename
had7bk.sen
pf, pm
Number of iterations 1000
Simulate variation in Biological parameters
TRUE
SR relationship constrained
TRUE
Ricker
910/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpr ADMB Beta Unscaled $f$ Unscaled E AIC |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 0.94 | 0.43 | 42.25 | 8.95 | 0.81 | 0.35 | 4.34 | 0.03 | 50.2 |
| Mean | 1.50 | 0.56 | 70.33 | 14.78 | 1.08 | 0.70 | 11.02 | 0.06 |  |
| 5\%ile | 0.59 | 0.28 | 13.26 | 4.09 | 0.62 | 0.09 | 2.83 | 0.01 |  |
| 25\%ile | 0.86 | 0.39 | 18.28 | 5.70 | 0.80 | 0.34 | 4.48 | 0.03 |  |
| 50\%ile | 1.20 | 0.49 | 24.95 | 7.41 | 0.98 | 0.64 | 6.98 | 0.06 |  |
| 75\%ile | 1.86 | 0.65 | 42.15 | 9.98 | 1.24 | 0.94 | 11.62 | 0.08 |  |
| 95\%ile | 3.48 | 1.04 | 155.53 | 28.68 | 1.83 | 1.56 | 30.00 | 0.14 |  |
| CV | 0.59 | 0.50 | 5.21 | 4.65 | 0.39 | 0.67 | 1.31 | 0.67 |  |

Beverton-Holt
920/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpr ADMB Beta Unscaled $f$ Unscaled E AIC |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 1.48 | 0.25 | 94.53 | 10.59 | 0.71 | 1.14 | 61.17 | 6.86 | 50.0 |
| Mean | 1.32 | 0.21 | 843.31 | 40.21 | 0.58 | 1.41 | 289.59 | 101.27 |  |
| 5\%ile | 0.49 | 0.13 | 41.91 | 4.97 | 0.07 | 0.92 | 36.92 | 1.22 |  |
| 25\%ile | 0.71 | 0.17 | 76.83 | 8.11 | 0.29 | 1.16 | 53.40 | 6.00 |  |
| 50\%ile | 0.99 | 0.21 | 134.64 | 12.03 | 0.56 | 1.35 | 77.88 | 16.22 |  |
| 75\%ile | 1.62 | 0.24 | 280.26 | 22.37 | 0.82 | 1.61 | 149.00 | 42.55 |  |
| 95\%ile | 3.46 | 0.32 | 1667.87 | 87.26 | 1.18 | 2.06 | 592.54 | 247.14 |  |
| CV | 0.68 | 0.30 | 5.99 | 4.82 | 0.62 | 0.26 | 8.04 | 9.01 |  |

Smooth hockeystick
978/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpr ADMB Beta Unscaled $f$ Unscaled E AIC |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 0.80 | 0.28 | 64.71 | 8.29 | 0.44 | 0.78 | 1.66 | 13.42 | 50.3 |
| Mean | 0.82 | 0.28 | 138.99 | 10.21 | 0.44 | 0.99 | 1.67 | 17.04 |  |
| 5\%ile | 0.46 | 0.16 | 32.77 | 4.95 | 0.27 | 0.63 | 1.04 | 10.82 |  |
| 25\%ile | 0.60 | 0.21 | 52.04 | 7.07 | 0.35 | 0.76 | 1.33 | 13.04 |  |
| 50\%ile | 0.73 | 0.26 | 69.92 | 9.24 | 0.42 | 0.95 | 1.61 | 16.26 |  |
| 75\%ile | 0.92 | 0.32 | 103.52 | 12.14 | 0.52 | 1.21 | 1.96 | 20.85 |  |
| 95\%ile | 1.48 | 0.47 | 252.67 | 18.90 | 0.65 | 1.44 | 2.47 | 24.82 |  |
| CV | 0.47 | 0.42 | 3.00 | 0.45 | 0.28 | 0.27 | 0.28 | 0.27 |  |


| Per recruit |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Bmsypr | MSYpr | Fpa | Flim |
| Determinis | 0.20 | 0.18 | 0.20 | 0.28 | 1.46 | 0.19 |  |  |
| Mean | 0.17 | 0.15 | 0.18 | 0.28 | 2.58 | 0.19 |  |  |
| 5\%ile | 0.09 | 0.07 | 0.10 | 0.16 | 0.72 | 0.11 |  |  |
| 25\%ile | 0.15 | 0.13 | 0.15 | 0.21 | 1.04 | 0.14 |  |  |
| 50\%ile | 0.18 | 0.15 | 0.18 | 0.26 | 1.33 | 0.18 |  |  |
| 75\%ile | 0.21 | 0.18 | 0.21 | 0.32 | 1.82 | 0.21 |  |  |
| 95\%ile | 0.26 | 0.22 | 0.28 | 0.47 | 4.41 | 0.30 |  |  |
| CV | 0.31 | 0.31 | 0.32 | 0.43 | 3.04 | 0.34 |  |  |

Per recruit (human consumption + discards - for comparison only)

|  | F35 | F40 |  | F01 |  | Fmax |  | Bmsypr |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | MSYpr | Fpa | Flim |  |  |  |  |  |
| Determinis | 0.20 | 0.18 | 0.25 | 0.40 | 0.91 | 0.27 | - |  |
| Mean | 0.19 | 0.17 | 0.25 | 0.42 | 1.56 | 0.27 |  |  |
| 5\%ile | 0.10 | 0.09 | 0.13 | 0.24 | 0.37 | 0.20 |  |  |
| 25\%ile | 0.17 | 0.15 | 0.21 | 0.35 | 0.59 | 0.24 |  |  |
| $50 \%$ ile | 0.20 | 0.17 | 0.25 | 0.41 | 0.80 | 0.27 |  |  |
| 75\%ile | 0.22 | 0.19 | 0.29 | 0.48 | 1.15 | 0.30 |  |  |
| 95\%ile | 0.26 | 0.22 | 0.35 | 0.61 | 2.77 | 0.37 |  |  |
| CV | 0.25 | 0.25 | 0.28 | 0.29 | 3.04 | 0.19 |  |  |



Figure 7.4.1. Length distributions of the landings of haddock in VIIb-k in 2009. All French fleets were combined; IRL OTB is the Irish otter trawl fleet; UK beam is the UK beam trawl fleet and UK trawl is all trawls except beam.


Figure 7.4.2. Length distributions of discards and the retained catch of haddock in VIIb-k in 2009. FR OT_CRU is the French otter trawl Nephrops fleet; FR OT_DEF is the French otter trawl gadoid+benthic fleet; IRL OTB is the Irish otter trawl fleet; all UK fleets were combined. Irish data were raised to total numbers, the length distributions of the landings (from port sampling) is given for comparison.


Figure7.4.3a. Numbers-at-age of Irish Discards of haddock in VIIb and VIIgj. The Irish discards in VIIgj were raised to international levels using effort as auxiliary variable.


Figure7.4.3b. Proportion of discards by age and year.


Figure7.4.4. Age composition of the landings (grey) and discards (white).



Figure 7.4.5a. Log standardised indices of tuning fleets by year. The IR7bGFS and IR7jGFS were not used in the assessment. See Stock Annex for a description of the fleets.

$\infty$ VactacN $\rightarrow 0$

Figure 7.4.5b. Log standardised indices of tuning fleets by year class. The IR7bGFS and IR7jGFS were not used in the assessment. See Stock Annex for a description of the fleets.


Figure 7.4.6. Survey indices of recruitment-at-age 0 , presented on a linear and logscale. The EVHOE and SAGFS were used as tuning fleets. The IR GFS fleets are presently not used.


Figure 7.4.7. Lpue of haddock and effort for Irish Otter trawl fleets, the French gadoid fleet and effort only for UK trawl (all trawl gears except beam trawl) fleet.


Figure 7.4.8. Log catchability regressions and residual plots of the tuning fleets used in the assessment.

fshk FR7fghGAD FR7fghjEVHOE IR7bjOT IR7gSAGFS

Figure7.4.9. Scaled weights of the tuning fleets used in the assessment.


Figure7.4.10. Retrospective XSA analysis.


Figure7.4.11. Stock summary plot.


Figure7.4.12. Fishing mortality and selectivity-at-age, the blue crosses represent the most recent year. F was separated into a landings and discards component using the proportion of the catch numbers that were discarded for each age and year. Selectivity was estimated by dividing the $F$ matrix by the catch Fbar 2-5 for each year.


Figure7.4.13. Sensitivity of the $F$ reference points to the number of years over which average $F$ and weights-at-age are calculated ( $\mathrm{F}_{\text {msy }}$ based on hockey stick model)


Figure7.4.14. Fitted stock-recruit relationships with 1000 MCMC re-samples. The left-hand plots show the deterministic fit (blue) as well as the confidence intervals from converged estimates of $\mathrm{F}_{\text {msy }}$ (red). Right-hand panels show the fits from the first 100 converged MCMC re-samples for illustration. The legends show the number of converged values for $\mathrm{F}_{\text {msy }}$ from $\mathbf{1 0 0 0}$ re-samples.


Figure 7.4.15. Estimates of F reference points and equilibrium yield and SSB against mortality using a Beverton and Holt recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical. Note that F2009 represents the landings component of $\mathrm{F}_{\text {bar }}$.


Figure 7.4.16. Estimates of F reference points and equilibrium yield and SSB against mortality using a hockey stick recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical. Note that F2009 represents the landings component of Fbar.


Figure 7.4.17. Estimates of F reference points and equilibrium yield and SSB against mortality using a Ricker recruitment model. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals of the converged estimates (red) and the right hand plots show the fit for the first 100 re-samples for illustration. The top two plots are identical. Note that F2009 represents the landings component of Fbar.


Figure 7.4.18. Fitted of F reference points and equilibrium yield and SSB from the ADMB srmsymc package. The left-hand plot illustrate the deterministic fit (blue) and confidence intervals (red) and the right hand plots show the fit for the first 100 iterations. The top two plots are identical. Note that F2009 represents the landings component of Fbar.

### 7.5 Nephrops in Division VIIb (Aran Grounds, FU17)

## Type of assessment in 2010

UWTV based assessment using WKNEPH 2009 protocol as described in the Stock Annex. This year long-term reference points have been examined for this stock. Further description on the background is presented in Section 7.5.2.

## ICES advice applicable to 2009

Exploitation boundaries in relation to precautionary limits/considerations
"The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to 2007. This corresponds to landings of no more than 900 tonnes for the Aran Grounds (FU 17)."

## ICES advice applicable to 2010

## June 2010:

"Advises on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Ratio for Nephrops fisheries should be less than the lower bound of $F_{0.1}$ ranges for similar stocks (8\%).This corresponds to landings of no more than $505 t$ for the Aran Grounds stock."

Advice was re-opened in November after the 2009 UWTV survey results were available.

## November 2010:

"ICES recommends that on the basis of exploitation boundaries in relation to high long-term yield and low risk of depletion of production potential that the Harvest Ratio for Nephrops fisheries should be less than the lower bound of Fo.1 ranges for similar stocks (8\%). This corresponds to landings of no more than $704 t$ for the Aran Grounds stock."

### 7.5.1 General

## Stock description and management units

The Aran Grounds Nephrops stock (FU17) covers ICES rectangles 34-35 D9-E0 within VIIb. This stock is included as part of the TAC Area VII Nephrops which includes the following stocks: Irish Sea East and West (FU14, FU15), Porcupine Bank (FU16), northwestern Irish Coast (FU18), southeastern and southwestern Irish Coast (FU19) and the Celtic Sea (FU20-22).


The TAC is set for Subarea VII which does not correspond to the stock area (FU 17 is shaded light yellow). There is no evidence that the individual functional units belong to the same stock. The 2010 TAC is $22432 \mathrm{t}, 9 \%$ less than the 2009 TAC. No FU17 specific restrictions in TAC apply thus, up to $100 \%$ of the Area VII TAC could, in theory be taken within FU17.

## Management applicable to 2009 and 2010

TAC in 2009

| Species: Norway lobster <br> Nephrops norvegicus | Zone:VII <br> (NEP/07.) |  |
| :--- | :---: | :--- |
| Spain | 1479 |  |
| France | 5994 |  |
| Ireland | 9091 |  |
| United Kingdom | 8086 |  |
| EC | 24650 |  |
| TAC | 24650 | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

TAC in 2010

| Species: | Norway lobster <br> Nephrops norvegicus | Zone:VII <br> (NEP/07.) |
| :--- | :--- | :--- | :--- |
| Spain | 1346 |  |
| France | 5455 |  |
| Ireland | 8273 |  |
| United Kingdom | 7358 |  |
| EU | 22432 | Analytical TAC |
| TAC | 22432 |  |

The MLS implemented by EC is set at 25 mm CL i.e. 8.5 cm total length and this regulation is applied by the Irish and UK fleets whereas a more restrictive regulation adopted by the French Producers' Organisations ( 35 mm CL i.e. 11.5 cm total length) is applied by the French trawlers.

## Ecosystem aspects

This section is detailed in Stock Annex.

## Fishery description

Since 1996 the Republic of Ireland fleet had over $99 \%$ of the landings from this FU. A description of the fleet is given in the Stock Annex. 37 Irish trawlers reported landings from this FU in 2009. This is about $32 \%$ decrease compared with the number of vessels reporting in 2008. However, only 16 of these vessels reported landings in excess of 10 t . The majority of these vessels are based in the port of Ros-a-Mhíl. Typical vessel length is $13-38 \mathrm{~m}$ and engine power ranges from $120-870 \mathrm{~kW}$. The majority of the landings are made with 80 mm mesh. Fishing trips usually last 3-7 days.

The majority of the landings come from the grounds to the west and southwest of the Aran Islands known as the 'back of the Arans ground' (See Stock Annex). The fishery on the Aran Grounds operates throughout the year, weather permitting with a seasonal trend (Figure 7.5).

## Fishery in 2009

The 2009 landings decreased by $32 \%$ from 2008 to 625 t . The decline in landings is mainly attributable to a decline in fishing effort. A decommissioning programme was in operation in Ireland during 2007 and 2008. Eight vessels active in the FU17 fishery were decommissioned. These vessels accounted for approximately $25 \%$ of the landings from the 2007-2008 period. In addition increasing fuel prices and poor market value for Nephrops impacted in the activity of the remaining fleet during 2009. Landings in Q4 accounted for $\sim 50 \%$ of the 2009 landings which is somewhat unusual and this explains the relatively low proportion of females in the landings (males normally dominate the landings in Q4 see Annex).

### 7.5.2 Data

Sampling of landings and discards resumed in 2008 after a break of two years (20062007) in the sampling programme. This break was due to non-cooperation with sampling by the fishing industry. Sampling levels in 2009 were good and are detailed in Section 2 (Table 2.1). Historical data availability and quality is reported in the Stock Annex (Section B).

## Landings

The reported landings time-series is shown in Figure 7.5.1and Table 7.5.1. The reported Irish landings from FU17 have fluctuated around 800 t in the recent years. There is concern about the accuracy of reported landings statistics for Nephrops by Irish vessels due to restrictive quotas and various misreporting practices. The introduction of sales notes and increased control and enforcement since 2007 should improve the accuracy of reported landings data. The TAC was increased in 2007 and 2008 this has led to an increase in reported landings and lpue.

## Commercial cpue

Effort data for this FU is available from 1995 for the Irish otter trawl Nephrops directed fleet. In 2009 this fleet accounted for $\sim 90 \%$ of the landings compared with an average of $70 \%$ over the time period. These data have not been standardised to take into account vessel or efficiency changes during the time period. Effort has declined between 2003-2006 then increased in 2007 to 2008 and declined again in 2009 (Table 7.5.2.). Landings per unit of effort (lpues) have been fluctuating around an average of $39 \mathrm{~kg} / \mathrm{hr}$. Lpue in 2009 was above average at $52 \mathrm{~kg} / \mathrm{hr}$ (Figure 7.5.2).

## Discarding

Before 2001 there was no discard sampling and it was thought that Nephrops discarding in this fishery was relatively low. Since 2001 discard rates have been estimated using unsorted catch and discards sampling (as described in the Stock Annex). Discard rates range between $14-24 \%$ of total catch by weight and $25-40 \%$ of total catch by number (Table 7.5.3). Discard rate of females tends to be higher due to the smaller average size and market reasons. There is no information on discard survival rate in this fishery ( $10 \%$ is assumed). No estimates of discards were available from 2006 and 2007 due to the non co-operation of the fishing industry with sampling programmes.

## Biological sampling

Sampling programme resumed in 2008 and since then coverage and intensity has been very good. The mean size of whole Nephrops ( $>35 \mathrm{~mm}$ ) in Irish landings has re-
mained stable between 1995 and 2000 for both sexes (Figure 7.5.3 and Table 7.5.4). The mean size of Nephrops in the catch has remained relatively stable since 2001.

The sex ratio in the landings is slightly male biased (Figure 7.5.4). The proportion of males is high in 2009 due an increased proportion of the landings taken in autumn (see Fishery in 2009).

There is no change to other biological parameter as described in the Annex.

## Abundance indices from UWTV surveys

Prior to the 2010 WG burrow counts for 2004 and 2005 surveys were verified and there were also minor revisions to 2002 to 2007 where the survey data was quality controlled. This verification and QC process resulted in some changes to historical abundance estimates although it did not change the overall perception in the trend in the time-series (See Lordan and Doyle, WD8). WKNEPH 2009 concluded that this survey could be used as an absolute index of abundance for this stock provided the bias (see text table below) was taken into account. This direct use of the survey is in lieu of alternative assessment approaches. These bias sources are not easily estimated and are largely based on expert opinion. In the Aran Grounds the largest source of perceived bias is the "edge effect".

|  |  | species |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU | Area | Edge effect detection rate |  |  |  |  |
| identification | Occupancy Cumulative bias |  |  |  |  |  |
| 17 | Aran | 1.35 | 0.9 | 1.05 | 1 | 1.3 |

The blanked krigged contour plot and posted point density data are shown in Figure 7.5.5. The krigged contours correspond very well to the observed data. In general the densities are higher towards the western side of the ground rather and there is a notable trend towards lower densities towards the east. Densities and abundance have fluctuated considerably of the time-series (e.g. $0.6-1.4$ burrows $/ \mathrm{m}^{2}$ ). The mean density in 2009 is approx $30 \%$ increase on 2008 it remains below the average of the time-series.

The summary statistics from this geostatistical analysis are given in Table 7.5.5 and plotted in Figure 7.5.6. The 2009 estimate of 718 million burrows is the third lowest to date but the estimates have fluctuated fairly widely to date since the survey commenced. The estimation variance of the survey as calculated by EVA is relatively low (CVs in the order $<5 \%$ ). Random stratified estimates are given for the smaller Slyne Head and Galway Bay grounds. Currently the spatial extent of these other grounds not well estimated. The size and contribution to landings of these grounds is small relative to the Aran grounds and these have not been taken into account in the overall abundance estimate or catch options.

As in previous years the relationship between commercial lpue in the autumn and spring commercial fishery and survey abundance was explored in Figure 7.5.7. The results also suggest that there is a negative relationship between survey abundance in June and lpue in the autumn and a weakly positive relationship with the fishery in the subsequent spring.

### 7.5.3 Assessment

## Review Group comments on the 2009 assessment

The RG agrees with the WG on the assessment used and considers the data the best available. With that said, the RG disagrees with ICES (advice for 2009) and the WG advice (for 2009) that the fishery is sustainable at current levels of effort. The 2008 survey estimates of burrows is the lowest in the time-series at $\sim 60 \%$ of the time-series average with a high current $F$ and has the highest harvest rate in the short time-series. Effort has increased and is at the highest level in the time-series, but lpue ( $\mathrm{kg} / \mathrm{hr}$ ) decreased slightly from 2007 to 2008. The slight decrease in lpue coupled with the lowest abundance estimate in the time-series could be an indicator that the TAC could be set lower to decrease F and effort. This may be difficult given the changes in the fishery (i.e. increases in effort being displaced from other areas and misreporting of landing statistical area). Unless there is an observed increased in burrows in the next survey, a precautionary approach should be taken. (the UWTV abundance in 2009 increased by $\sim 30 \%$ ). Advice was reopened in November 2009 following the UWTV survey which showed that abundance in 2009 had increased by ~30\%.

## Approach in 2010

The assessment approach used by WGCSE 2010 is consistent with that set out in the Stock Annex and WKNEPH (2009). Exploratory SCAs (Separable cohort analysis) were carried out to derive suitable reference points for this stock. These SCAs used 2008 and 2009 sampling data and combined 2008-2009 sampling. Different selection patterns between sexes were included in the model to take into account differences in selection observed in the fishery.

## Comparison with previous assessments

The assessment is based on similar methods and data as used in 2009. The stock size is estimated to have increased and harvest ratio has decreased based on the UWTV survey.

## State of the stock

UWTV abundance estimates suggest that the stock size has fluctuated widely. A diffusion model (Dennis et al., 1991) that takes into account stochastic variability in the estimates results in a slope or $\mu$ close to 0 indicating no overall trend in abundance. The fluctuations in survey abundance appear to be largely independent of harvest rate estimates from the fishery. This may suggest that natural or other unaccounted mortality and recruitment are largely responsible for the observed variability (assuming landings and UWTV abundance estimates are accurate).

Table 7.5.6 summarises recent harvest ratios for the stock along with other stock parameter. Figure 7.5 .8 is the stock summary plot for FU17. Recent harvest rates have fluctuated around $8 \%$, abundance has fluctuated around 600 million and landing have fluctuated around 800 t .

### 7.5.4 Short-term projections

Forecast inputs and historical estimates of mean weight in landings and harvest ratios are presented in Table 7.5.6. Since 2002 mean weight in the landings has varied between $18-27 \mathrm{grs}$. The estimate harvest ratio has also varied a lot, $3-13 \%$ with 2008 being the highest observed.

A prediction of landings for 2011 was made for the Aran Grounds Functional Unit using the approach agreed procedure proposed at WKNEPH 2009 and outlined in the Stock Annex. The notable difference is that for FU17 a two year average landing weight and proportion retained by the fishery has been used due to lack of sampling in 2007-2008. Table 7.5 .7 shows landings predictions at various harvest ratios, including those equivalent to fishing within the range of $\mathrm{F}_{0.1}$ to $\mathrm{F}_{\text {max. }}$. The 2009 harvest ratio for the Aran grounds is estimated to be below $\mathrm{F}_{\text {max }}$.

### 7.5.5 MSY explorations

As discussed in Section 2.2 no dynamic population model is fitted to the data so no estimates of spawning stock and recruitment were available to determine $\mathrm{F}_{\text {msy }}$. In response to the recommendations of WKFRAME (2010), the Bell/Dobby combined sexlength cohort analysis (SCA) model used to determine Harvest Rates associated with fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ at WKNEPH (2009) was adapted to also output estimates of $\mathrm{F}_{35 \% \mathrm{Spr}}$ and to take into account separate sex selection. The SCA model fits are presented in Figure 7.5.9. These F estimates could be used as a proxy for $\mathrm{F}_{\mathrm{msy}}$. Catchlength data were available for the most recent two years in the fishery 2008 and 2009. For other stocks three years of length data were used in the analysis but in this case there was a gap in sampling in 2006 and 2007.

The results of the model in the text table below show the F multipliers required to achieve the potential $\mathrm{F}_{\mathrm{msy}}$ proxies, the harvest rates that correspond to those multipliers and the resulting level of spawner per recruit as a percentage of the virgin level.

|  |  | Fbar 20-40mm |  | Harvest Rate | \% Virgin Spawner per Recruit |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: |
|  | Female | Male |  | Female |  |  |
| $\mathrm{F}_{0.1}$ | Comb |  | 0.06 | 0.17 | Male |  |
| $\mathrm{F}_{0.1}$ | Female | 0.11 | 0.31 | $7.2 \%$ | $64.3 \%$ |  |
| $\mathrm{~F}_{0.1}$ | Male | 0.05 | 0.14 | $9.1 \%$ | $49.7 \%$ |  |
| $\mathrm{~F}_{35}$ | Comb | 0.12 | 0.34 | $6.4 \%$ | $68.8 \%$ |  |
| $\mathrm{~F}_{35 \%}$ | Female | 0.55 | 0.19 | $10.5 \%$ | $47.0 \%$ |  |
| $\mathrm{~F}_{35 \%}$ | Male | 0.07 | 0.21 | $12.8 \%$ | $34.9 \%$ |  |
| $\mathrm{~F}_{\max }$ | Comb | 0.12 | 0.34 | $8.4 \%$ | $60.0 \%$ |  |
| $\mathrm{~F}_{\max }$ | Female | 0.56 | 0.19 | $11.1 \%$ | $47.0 \%$ |  |
| $\mathrm{~F}_{\max }$ | Male | 0.09 | 0.26 | $13.0 \%$ | $34.5 \%$ |  |
|  |  |  |  | $9.8 \%$ | $54.1 \%$ |  |

This fishery is highly seasonal (see Annex), but the timing of the fishery has varied somewhat in recent years. In 2009 a larger proportion of the landings were taken in autumn leading to a change in sex ratio and size compared with 2008. This coupled with limited time-series of survey data and biological knowledge of the stock suggest a risk adverse harvest rate would be appropriate.

Compared to other Nephrops fisheries in ICES area the absolute population density of this stock is relatively high Figure 6.5.9. This implies that sperm limitation if males are over fished is not likely to be a significant problem. The combined sex $\mathrm{F}_{35 \%}$ SPR would result in $>20 \%$ males SPR and $47 \%$ female SPR. This combined sex $\mathrm{F}_{35} \%$ also corresponds to $\mathrm{F}_{\text {max. }}$. The WG concluded that a combined sex $\mathrm{F}_{35} \%$ was a suitable $\mathrm{F}_{\text {msy }}$ proxy for this stock. This corresponds to a harvest rate of $9.7 \%$.

### 7.5.6 Biological reference points

Precautionary reference points have not been defined for Nephrops stocks. Given the short time-series of UWTV survey data it is not possible to define an appropriate $B_{\text {trig- }}$ ger. $\mathrm{F}_{35 \%}$ SPR is proposed by the WG as proxy for $\mathrm{F}_{\text {msy }}$.

### 7.5.7 Management strategies

As yet there are no explicit management strategies for this stock but there have been some discussions amongst the fishing industry and scientists about developing a long-term plan for the management of the Aran fishery. Sustainable utilisation of the Nephrops stock will form the cornerstone of any management strategy for this fishery.

### 7.5.8 Uncertainties and bias in assessment and forecast

There are several key uncertainties and bias sources in the method proposed (these are discussed further in WKNEPH 2009). Various agreed procedures have been put in place to ensure the quality and consistency of the survey estimates following the recommendations of several ICES groups (WKNEPTV 2007; WKNEPHBID 2008; SGNEPS 2009). These recommendations have been retrospectively applied to historical survey estimates this year (Section 5.1) and these are now considered final. Taking explicit note of the likely biases in the surveys may at least provide an estimate of absolute abundance that was more accurate but no more precise (WKNEPH 2009). The survey estimates themselves are likely to be fairly precisely estimated given the homogeneous distribution of burrow density and the modelling of spatial structuring. The cumulative bias estimates for FU17 are largely based on expert opinion. The precision of these cannot yet be characterised. Ultimately there still remains a degree of subjectivity in the production of UWTV indices.

In the provision of catch options based on the absolute survey estimates additional uncertainties related to mean weight in the landings and the discard rates also arise. For FU17 deterministic estimates of the mean weight in the landings and discard rates for 2008 and 2009 have been used since sampling data was not available for the previous two years. Historical data suggest parameters have been variable in the past (Table 7.5.6). In future years the uncertainty in these key parameters should be estimated.

Landings data are assumed to be accurate. Since 2007 the introduction of "buyers and sellers legislation" in Ireland is thought to have improved the accuracy of the reported landings.

Finally, the catch options developed do not have any additional catches for the smaller Slyne or Galway Bay Grounds. This is likely to cause a small ( $<3 \%$ ) underestimate in the catch options for FU17 as a whole.

### 7.5.9 Recommendation for next Benchmark

This stock was benchmarked in 2009. WKNEPH 2009 suggested several areas to be addressed before the next Benchmark. Currently there is no recommended time frame for another benchmark.

### 7.5.10 Management considerations

The trends from the fishery (landings, effort lpue, mean size, etc.) appear to be relatively stable. Lpues have been relatively high in the last three years. Conversely, the UWTV abundance and mean density estimates show large fluctuations in burrow
abundance and harvest rates. This suggests that the Nephrops population at current exploitation and recruitment rates is rather dynamic. The generally low apparent harvest rate ( $9 \%$ average) appears to have little impact on observed stock fluctuations. A new survey point should be available after June 2010 which will provide a more up to date prognosis of stock status. The use of the most up to date survey information should be considered for this stock.

In recent years several newer vessels specialising in Nephrops fishing have participated in this fishery. These vessels target Nephrops on several other grounds within the TAC area and move around to optimise catch rates. Since the introduction of effort management associated with the cod long-term plan (EC 1342/2008) there have been concerns that effort will be displaced towards the Aran and other Nephrops grounds where effort control has not been put in place. This has not occurred in 2009 and effort is down substantially ( $-37 \%$ ). This can be explained by the decommissioning of several vessels that actively participated in the fishery heretofore, and generally poor economic conditions for this fishery. Nevertheless management measures should be established to prevent unsustainable increases in effort and catch for this stock.

### 7.5.11 References

Dennis, B., P.L. Munholland and J.M. Scott. 1991. Estimation of growth and extinction parameters for endangered species. Ecological Monographs 64:205-224.

Table 7.5.1 Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.

|  | FU 17 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | France | Rep. of <br> Ireland | UK | Total |  |
| 1974 | 477 |  |  | 477 |  |
| 1975 | 822 |  |  | 822 |  |
| 1976 | 131 |  |  | 131 |  |
| 1977 | 272 |  |  | 272 |  |
| 1978 | 481 |  |  | 481 |  |
| 1979 | 452 |  |  | 452 |  |
| 1980 | 442 |  |  | 442 |  |
| 1981 | 414 |  |  | 414 |  |
| 1982 | 210 |  |  | 210 |  |
| 1983 | 131 |  |  | 131 |  |
| 1984 | 324 |  |  | 324 |  |
| 1985 | 207 |  |  | 207 |  |
| 1986 | 147 |  |  | 147 |  |
| 1987 | 62 | 814 |  | 62 |  |
| 1988 | 14 | 317 |  | 828 |  |
| 1989 | 27 | 389 |  | 344 |  |
| 1990 | 30 | 439 |  | 519 |  |
| 1991 | 11 | 399 |  | 372 |  |
| 1992 | 11 | 361 | 361 | 0 |  |
| 1993 | 11 | 707 | 4 | 372 |  |
| 1994 | 18 | 774 | 1 | 729 |  |
| 1995 | 91 | 519 | 4 | 866 |  |
| 1996 | 2 | 839 | 0 | 525 |  |
| 1997 | 2 | 1401 | 0 | 1410 |  |
| 1998 | 9 | 1140 | 0 | 1140 |  |
| 1999 | 0 | 879 | 0 | 880 |  |
| 2000 | 1 | 912 | 0 | 913 |  |
| 2001 | 1 | 1152 | 0 | 1154 |  |
| 2002 | 2 | 933 | 0 | 933 |  |
| 2003 | 0 | 525 | 0 | 525 |  |
| 2004 | 0 | 778 | 0 | 778 |  |
| 2005 | 0 | 637 | 0 | 637 |  |
| 2006 | 0 | 913 | 0 | 913 |  |
| 2007 | 0 | 1050 | 7 | 1057 |  |
| 2008 | 0 | 625 | 0 | 625 |  |
| 2009 | 0 |  |  |  |  |
|  |  |  |  |  |  |

Table 7.5.2. Nephrops in FU 17 (Aran Grounds). Irish effort and lpue for Nephrops directed fleet.

| Year | Irish Nephrops Directed Fleet |  |  |
| :---: | :---: | :---: | :---: |
|  | Effort (Hrs) | Landings (tonnes) | LPUE (kg/hr) |
| 1995 | 15306 | 530 | 34.6 |
| 1996 | 9109 | 311 | 34.1 |
| 1997 | 15763 | 478 | 30.3 |
| 1998 | 21909 | 926 | 42.3 |
| 1999 | 19546 | 743 | 38.0 |
| 2000 | 17131 | 547 | 31.9 |
| 2001 | 18700 | 600 | 32.1 |
| 2002 | 18565 | 861 | 46.4 |
| 2003 | 19922 | 732 | 36.8 |
| 2004 | 12899 | 381 | 29.5 |
| 2005 | 14900 | 729 | 45.8 |
| 2006 | 10798 | 559 | 51.8 |
| 2007 | 13608 | 815 | 59.9 |
| 2008 | 16676 | 963 | 57.8 |
| 2009 | 10620 | 561 | 52.8 |

Table 7.5.3. Nephrops in FU17 (Aran Grounds). Landings and discard weight and numbers by year and sex.

|  | Female |  | Male |  | Both sexes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Discards (t) | Landings (t) | Discards (t) | \% Discard |
| 2001 | 312 | 109 | 601 | 138 | 21\% |
| 2002 | 423 | 96 | 729 | 99 | 14\% |
| 2003 | 237 | 89 | 688 | 98 | 17\% |
| 2004 | 267 | 71 | 259 | 45 | 18\% |
| 2005 | 323 | 106 | 441 | 86 | 20\% |
| 2006 2007 |  |  | No Sampling |  |  |
| 2008 | 324 | 160 | 726 | 98 | 20\% |
| 2009 | 90 | 130 | 726 | 134 | 24\% |


|  | Female Numbers '000s |  | Male Numbers '000s |  | Both sexes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Discards | Landings | Discards | \% Discard |
| 2001 | 18,665 | 12,161 | 29,949 | 13,250 | $34 \%$ |
| 2002 | 23,105 | 9,374 | 31,256 | 8,326 | $25 \%$ |
| 2003 | 14,530 | 9,577 | 29,538 | 8,744 | $29 \%$ |
| 2004 | 16,109 | 7,068 | 12,930 | 4,282 | $28 \%$ |
| 2005 | 20,280 | 11,383 | 21,828 | 8,967 | $33 \%$ |
| 2006 | No Sampling |  |  |  |  |
| 2007 | 15,697 | 13,223 | 31,184 | 8,350 | $32 \%$ |
| 2008 | 15,085 | 20,421 | 8,218 | $40 \%$ |  |

Table 7.5.4. Nephrops in FU17 (Aran Grounds). Mean size trends for catches and whole landings by sex.

| Year | Catches |  | Catches |  | Whole Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | <35mm CL |  | >35mm CL |  | <35mm CL |  | >35mm CL |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 1995 | na | na | na | na | 32.0 | 31.8 | 38.3 | 37.0 |
| 1996 | na | na | na | na | 31.1 | 32.1 | 37.8 | 37.4 |
| 1997 | na | na | na | na | 31.9 | 32.0 | 37.8 | 37.4 |
| 1998 | na | na | na | na | 31.3 | 31.7 | 38.0 | 37.2 |
| 1999 | na | na | na | na | 31.3 | 32.3 | 38.0 | 37.1 |
| 2000 | na | na | na | na | 32.0 | 31.4 | 38.4 | 36.3 |
| 2001 | 28.9 | 27.5 | 38.0 | 37.3 | na | na | na | na |
| 2002 | 30.7 | 29.1 | 38.2 | 37.2 | na | na | na | na |
| 2003 | 30.5 | 27.4 | 38.2 | 38.0 | na | na | na | na |
| 2004 | 29.3 | 28.3 | 37.3 | 37.5 | na | na | na | na |
| 2005 | 28.9 | 27.7 | 37.8 | 37.2 | na | na | na | na |
| 2006 | No Sampling |  |  |  |  |  |  |  |
| 2008 | 27.4 | 29.7 | 36.8 | 37.8 | na | na | na | na |
| 2009 | 30.3 | 28.4 | 38.0 | 37.1 | na | na | na | na |

## na $=$ not available

Table 7.5.5. Nephrops in FU17 (Aran Grounds). Results summary table for geostatistical analysis of UWTV survey.

| Ground | Year |  | Mean Density (No.IM2) | $\begin{gathered} \text { Domain Area } \\ (\mathrm{m} 2) \end{gathered}$ | Geostatistical abundance estimate (million burrows) | CV on Burrow Estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Aran | 2002 | 49 | 0.84 | 943 | 818 | 4\% |
|  | 2003 | 41 | 1.01 | 943 | 989 | 5\% |
|  | 2004 | 64 | 1.43 | 943 | 1397 | 3\% |
|  | 2005 | 70 | 1.09 | 936 | 1063 | 3\% |
|  | 2006 | 67 | 0.64 | 932 | 621 | 3\% |
|  | 2007 | 71 | 0.93 | 942 | 906 | 3\% |
|  | 2008 | 62 | 0.57 | 842 | 515 | 3\% |
|  | 2009 | 82 | 0.73 | 940 | 718 | 2\% |

Results summary table for empirical statistical analysis of UWTV survey

| Ground | Year | Number of stations | Mean Density (No./M ${ }^{2}$ ) | Area Surveyed (m2) | Burrow count | Standard Deviation | 95\%CI | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Galway Bay | 2002 | 7 | 1.58 | 1,299 | 2,017 | 0.37 | 0.34 | 9\% |
|  | 2003 | 3 | 1.60 | 591 | 941 | 0.29 | 0.73 | 11\% |
|  | 2004 | 9 | 0.73 | 2,312 |  | 0.42 | 0.32 | 19\% |
|  | 2005 | 4 | 1.67 | 661 | 1,625 | 0.20 | 0.32 | 6\% |
|  | 2006 | 3 | 0.98 | 540 | 1,107 | 0.27 | 0.67 | 16\% |
|  | 2007 | 5 | 1.14 | 890 | 992 | 0.24 | 0.29 | 9\% |
|  | 2008 | 10 | 0.42 | 1,907 | 859 | 0.31 | 0.22 | 23\% |
|  | 2009 | 8 | 0.93 | 1,207 | 1,116 | 0.16 | 0.14 | 6\% |
| Slyne Grounds | 2002 | 5 | 0.85 | 1,216 | 1,027 | 0.19 | 0.23 | 10\% |
|  | 2003 | 0 | - | - | - | - | - | - |
|  | 2004 | 3 | 0.68 | 827 | 531 | 0.27 | 0.66 | 23\% |
|  | 2005 | 3 | 0.55 | 531 | 294 | 0.05 | 0.13 | 6\% |
|  | 2006 | 3 | 0.41 | 526 | 210 | 0.20 | 0.49 | 28\% |
|  | 2007 | 4 | 0.63 | 838 | 547 | 0.31 | 0.49 | 24\% |
|  | 2008 | 0 | - | - | - | - | - | - |
|  | 2009 | 6 | 0.40 | 531 | 144 | 0.22 | 0.23 | 22\% |

[^14]Table 7.5.6. Nephrops in FU17 (Aran Grounds). Forecast inputs (highlighted) and historical estimates of mean weight in landings and harvest ratio.

| Year | Landings in Number (millions) | Discards <br> in <br> Number <br> (millions) | Removals <br> in <br> Number <br> (millions) | Prop <br> Removals <br> Retained | Adjusted <br> Survey <br> (millions) | Harvest <br> Ratio | Landings (t) | Discards <br> (t) | Mean <br> Weight <br> in <br> landings <br> (gr) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 48.7 | 25.4 | 71.6 | 0.68 |  |  | 913 |  |  |
| 2002 | 54.5 | 17.7 | 70.4 | 0.77 | 629 | 11.2\% | 1,154 | 192 | 21.2 |
| 2003 | 44.1 | 18.3 | 60.6 | 0.73 | 761 | 8.0\% | 933 | 183 | 21.2 |
| 2004 | 29.0 | 11.4 | 39.3 | 0.74 | 1075 | 3.7\% | 525 | 112 | 18.1 |
| 2005 | 42.4 | 19.7 | 60.1 | 0.70 | 818 | 7.4\% | 778 | 182 | 18.4 |
| 2006 | na | na | 49.5 | na | 478 | 10.4\% | 637 | na | na |
| 2007 | na | na | 57.3 | na | 697 | 8.2\% | 913 | na | na |
| 2008 | 46.9 | 21.6 | 66.3 | 0.71 | 396 | 16.7\% | 1,057 | 245 | 22.5 |
| 2009 | 23.5 | 15.7 | 37.6 | 0.62 | 552 | 6.8\% | 625 | 256 | 26.6 |
| Avg 08 \&09 |  |  |  | 0.67 |  |  |  |  | 24.6 |

na= not available due to non-cooperation with sampling programmes.
Shading indicates removal estimated based on combined 2005 and 2008 numbers-at-length scaled appropriately to landings in 2006 and 2007. The commensurate harvest ratio estimate is also shaded.

Table 7.5.7. Nephrops in FU 17 (Aran Grounds). Short-term forecast management option table giving catch options for 2011.

|  |  | Implied fishery |  |  |
| :--- | :---: | :---: | :---: | :---: |
|  | Harvest rate | Survey Index <br> (millions) | Retained <br> number <br> (millions) | Landings (tonnes) |
| MSY framework | $10.5 \%$ | 552 | 39 | 948 |
| F 2010 (avg. 2007-2009) | $10.6 \%$ | 552 | 39 | 957 |
| F0.1 Combined | $7.2 \%$ | 552 | 26 | 650 |
| Fmax Combined | $11.1 \%$ | 552 | 41 | 1,002 |
| F0.1 Comb | $7.2 \%$ | 552 | 26 | 650 |
| F0.1 Female | $9.1 \%$ | 552 | 33 | 822 |
| F0.1 Male | $6.4 \%$ | 552 | 24 | 578 |
| F35\% Comb | $10.5 \%$ | 552 | 39 | 948 |
| F35\% Female | $12.8 \%$ | 552 | 47 | 1,156 |
| F35\% Male | $8.4 \%$ | 552 | 31 | 759 |
| Fmax Comb | $11.1 \%$ | 552 | 41 | 1,002 |
| Fmax Female | $13.0 \%$ | 552 | 48 | 1,174 |
| Fmax Male | $9.8 \%$ | 552 | 36 | 885 |
|  | $2.0 \%$ | 552 | 7 | 181 |
|  | $4.0 \%$ | 552 | 15 | 361 |
|  | $6.0 \%$ | 552 | 22 | 542 |
|  | $5.0 \%$ | 552 | 29 | 722 |
|  | $10.0 \%$ | 552 | 37 | 903 |
| Landings Mean Weight (Kg) |  | 552 | 44 | 1,084 |
| Survey Overestimate Bias |  | 552 | 51 | 1,264 |
| Survey Numbers (Millions) |  | $12.0 \%$ | 718 |  |
| Prop. Retained by the Fishery |  | 0.67 |  | Basis |
|  | $14.0 \%$ |  |  | Sampling 2008 and 2009 |
|  |  |  |  | UKNEPH 2009 |
|  |  |  |  |  |



Figure 7.5.1. Nephrops in FU17 (Aran Grounds). Landings in tonnes by country.


Figure 7.5.2. Nephrops FU17 Aran Grounds. Irish effort and lpue for Nephrops directed fleet.

## Length frequencies for catch (dotted) and landed(solid): Nephrops in FU17


mis (25mm) and 33mm levels displayed

Figure 7.5.3. Nephrops FU17 Aran Grounds. Length distributions in the catches 2001-2005, 20082009 and in the landings 1995-2001.


Figure 7.5.4. Nephrops in FU17 (Aran Grounds). Sex ratio of whole landings (1995-2000), landings (2001-2009) and catch (2001-2009).


Figure 7.5.5. Nephrops in FU17 (Aran Grounds). Contour plots of the krigged density estimates for the Aran Ground UWTV surveys from 2002-2009.


Figure 7.5.6. Nephrops FU17 Aran Grounds.
a)

b)



Figure 7.5.7. Nephrops FU 17 Aran Grounds.
a) The monthly lpue from FU17 (bars) and survey abundance index (red dots).
b) mean standardised long-term (1995-2009) seasonal trend in lpue for FU17.
c) the relationship between lpues for two time periods and survey abundance estimates.


FU 17 : Harvest rate


Figure 7.5.8. Nephrops FU17 Aran Grounds. Stock Summary plots: Landings (tonnes), UWTV abundance (millions) and Harvest Ratio (\% dead removed/UWTV abundance).


Figure 7.5.9. Nephrops FU17 Aran Grounds.SCA outputs based on 2008 and 2009 length distributions.

### 7.6 Nephrops in Division VIIb, c,j,k (Porcupine Bank, FU 16 )

Type of assessment in 2010
This year the Working Group updated the fishery information, survey data and other indicators for Nephrops in Division VIIbcjk. There are recruits appearing in the survey and commercial catches for the first time in several years. The landings in recent years have been dominated by increasing larger individuals and higher proportions of females. All indicators suggested that the stock was over exploited and on the point of collapse. The new recruitment offers an opportunity to begin the rebuild of the stock if exploitation rates can be kept low.

## ICES advice applicable to 2009

"Because of the apparent low recruitment and the recent expansion of the fishery there is an associated increased exploitation. ICES recommends reduction in the exploitation rate and restricting catches in 2009 to no more than $1000 t$, which corresponds to the catch level before the expansion of the fishery (2000-2003). The fishery should not be allowed to expand again unless it can be shown that it is sustainable".

## ICES advice applicable to 2010

"ICES advises on the basis of exploitation boundaries in relation to precautionary considerations that catches in 2010 should be reduced to the lowest possible level."

### 7.6.1 General

## Stock description and management units

The TAC area is Subarea VII. The Functional Unit for assessment includes some parts of the following ICES Divisions VIIb, $c$, $\mathfrak{j}$, and $k$. The exact stock area is shown on the map below and includes the following ICES Statistical rectangles: 31-35 D5-D6; 3235 D7-D8.


The TAC is set for all of Subarea VII. The FU16 is shaded light yellow and the closed area from 01/05/10-31/07/10 is within the blue dotted line.

## Management applicable to 2009 and 2010

TAC in 2009

| Species: Norway lobster <br> Nephrops norvegiaus | Zone:VII <br> (NEP/07.) |  |
| :--- | :---: | :--- |
| Spain | 1479 |  |
| France | 5994 |  |
| Ireland | 9091 |  |
| United Kingdom | 8086 |  |
| EC | 24650 |  |
| TAC | 24650 | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

TAC in 2010

| Species: | Norway lobster <br> Nephrops norvegicus | Zone:VII <br> (NEP/07.) |  |
| :--- | :--- | :--- | :--- |
| Spain | 1346 |  |  |
| France | 5455 |  |  |
| Ireland | 8273 |  |  |
| United Kingdom | 7358 |  |  |
| EU | 22432 | Analytical TAC |  |
| TAC | 22432 |  |  |

## Closed area restrictions

A closed area is in operation in 2010 (Council Regulation 23/2010 as of 14 January 2010 fixing for 2010 the fishing opportunities for certain fish stocks and groups of fish stocks). The closed area is shown in the map above.

## Article 11

## Restrictions on the use of certain fishing opportunities

During the period from 1 May to 31 July 2010 it shall be prohibited to fish for or retain on board any marine organisms other than herring, mackerel, pilchard/sardines, horse mackerel, sprat, blue whiting and argentines within the area bounded by rhumb lines sequentially joining the following positions:

| Point | Latitude | Longitude |
| :--- | :--- | :--- |
| 1 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 19^{\prime} \mathrm{W}$ |
| 2 | $52^{\circ} 40^{\prime} \mathrm{N}$ | $12^{\circ} 30^{\prime} \mathrm{W}$ |
| 3 | $52^{\circ} 47^{\prime} \mathrm{N}$ | $12^{\circ} 39,600^{\prime} \mathrm{W}$ |
| 4 | $52^{\circ} 47^{\prime} \mathrm{N}$ | $12^{\circ} 56^{\prime} \mathrm{W}$ |
| 5 | $52^{\circ} 13,5^{\prime} \mathrm{N}$ | $13^{\circ} 53,830^{\prime} \mathrm{W}$ |
| 6 | $51^{\circ} 22^{\prime} \mathrm{N}$ | $14^{\circ} 24^{\prime} \mathrm{W}$ |
| 7 | $51^{\circ} 22^{\prime} \mathrm{N}$ | $14^{\circ} 03^{\prime} \mathrm{W}$ |
| 8 | $52^{\circ} 10^{\prime} \mathrm{N}$ | $13^{\circ} 25^{\prime} \mathrm{W}$ |
| 9 | $52^{\circ} 32^{\prime} \mathrm{N}$ | $13^{\circ} 07,500^{\prime} \mathrm{W}$ |
| 10 | $52^{\circ} 43^{\prime} \mathrm{N}$ | $12^{\circ} 55^{\prime} \mathrm{W}$ |
| 11 | $52^{\circ} 43^{\prime} \mathrm{N}$ | $12^{\circ} 43^{\prime} \mathrm{W}$ |
| 12 | $52^{\circ} 38,800^{\prime} \mathrm{N}$ | $12^{\circ} 37^{\prime} \mathrm{W}$ |
| 13 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 23^{\prime} \mathrm{W}$ |
| 14 | $52^{\circ} 27^{\prime} \mathrm{N}$ | $12^{\circ} 19^{\prime} \mathrm{W}$ |

The following TCMs are in place for Nephrops in VII (excluding VIIa) after EC 850/9 in operation since 2000:

Minimum Landing Sizes (MLS); total length $>85 \mathrm{~mm}$, carapace length $>25 \mathrm{~mm}$, tail length $>46 \mathrm{~mm}$. Although it is legal to land smaller prawns from this fishery, marketing restrictions imposed by producer organisations in France mean smaller Nephrops ( $<35 \mathrm{~mm}$ CL or 115 mm whole length) are not retained in this fishery.

The mesh size restrictions apply to towed gears in VIIb-k targeting Nephrops and are given in Section 7.1. Vessels mainly used $80-99 \mathrm{~mm}$ mesh to target Nephrops on the Porcupine Bank.

## Fishery in 2009

The Nephrops fisheries in this area are very seasonal and rather sporadic, mainly targeting Nephrops when available and when weather conditions are good. At other times the vessels switch to other fisheries. Effort by French and Irish vessels in 2009 reduced substantially ( $>40 \%$ ) compared with 2008. Total international landings (Figure 7.6.1 and Table 7.6.1) in 2009 were the second lowest observed and of similar magnitude to landings in 2000 and 2003.

## Effect of regulations

Landings for the TAC area (Subarea VII) are undershot (Table 7.8.4). UK and Irish national quotas are restrictive but uptake by France and Spain is well below their quotas due to changes in relative landings from different FUs within this TAC area (Section 7.1). In the past TACs and quotas applied to the whole of VII do little to restrict the FU16 fishery. The closed area to be implemented in 2010 is coincident with a time period where the majority of annual international landings have been taken (see text table below). It is also spatial coincident with the main fishery (Figure 7.6.10). It is therefore expected to be quite effective at reducing fishing mortality provided that effort is not increase outside the time/area to compensate for the closure.

|  | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2003-2008 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\%$ of annual Int. landings taken May-July | $60 \%$ | $53 \%$ | $64 \%$ | $54 \%$ | $67 \%$ | $68 \%$ | $61 \%$ |

There has been discarding of small and maturing female Nephrops in this fishery in the past due to lower market price but there are no reliable estimates of this during the time-series. Discarding patterns are known to vary between countries.

### 7.6.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1.
Length compositions of annual landings are available from Spain (1986-2009), France (1995-2007) and Ireland (1995-2005 and 2008-2009). No sampling was possible in 2006 and 2007 for Ireland due to the withdrawal of co-operation with scientific sampling programmes by the fishing industry. Sampling in Ireland resumed in 2008. There was no sampling in France in 2008 and 2009 due to low landings. Sampling intensity in Spain was extremely low in 2008 and 2009 (two and five samples).

Sampling of Nephrops in this area is hampered by several factors:

- The remote nature of the fishery.
- Trips are long duration sometimes fishing in multiple areas.
- An increasing proportion of the landings are landed frozen or graded at sea.
- There is reluctance from fishermen and processors to allow sampling of landings due to high value of the larger Nephrops and the risk of damage to individuals during sampling.

These issues need to be resolved as current sampling intensity is insufficient to get precise and accurate length structure data of the catches.

Despite the low sampling intensity in recent years, the trends in indicators such as length and sex ratio are consistent across all countries and in the survey.

## Landings

Data on the mean size (carapace length, CL) of male and female Nephrops in the landings are available from Spain, France and Ireland (Table 7.6.2, Figure 7.6.2). The longest time-series are from Spain and, prior to 2002, these have been quite stable at between 39 and 43 mm CL for the males, and between 34 and 38 mm CL for the females. Since 2002 there has been an increasing trend in the mean size in the landings peaking in 2008. Mean Nephrops sizes in French landings also show an increasing
trend in both sexes. Mean sizes in the landings of Irish trawlers are more variable but clearly show increasing trend over the last number of years.
Raised frequency distributions of the sampled landings by sex are given in Figure 7.6.3. This also shows significant shift towards larger individuals in the landings since 2002 and few individuals at smaller sizes. The 2009 data for males shows a recruiting year class entering the landings at $\sim 35 \mathrm{~mm}$ CL. This is the first time in the time series a very obvious year-class signal has appeared in the landings-length distributions (though there are possibly other YC appearing at a slightly large size in other years).

It is difficult to extract other useful signals in the length frequency distributions plot, so for males a number of indicators were calculated (Figure 7.6.4). These included a recruitment proxy ( $\%$ of males $<32 \mathrm{~mm} \mathrm{CL}$ ), and percentage of larger individuals ( $>50 \mathrm{~mm} \mathrm{CL}$ ) in the sampled landings. An exploitation proxy was calculated using the slope of $\ln (C L)$ versus $\ln$ (Numbers) between $41-56 \mathrm{~mm}$ CL i.e. the slope of downward limb on the Right-Hand-Side of the length frequency distribution.

These indicators suggest the following: recruitment has fluctuated in the past and recruitment in the last five years (2004 to 2008) has probably been very weak. Recruitment in 2009 has more average levels (note: this conclusion is relatively insensitive to length threshold). The fishery in recent years exploits a higher proportion of larger individuals than ever before in the time-series. The exploitation proxy shows an increasing trend (i.e. steepness) since the early 2000s. The exploitation proxy in 2009 remains among the highest in the series.

## Discards

There are no estimates of discards for this stock but the intra-country differences in size structure of the landings suggest different on-board selection patterns (mesh sizes used are broadly similar across fleets).

## Biological

In the most recent years there has been a large change in sex ratio in the landings and survey catches (Figure 7.6.5). See section below for survey details. The change in sex ratio in the landings is strongly influenced by the re-availability of data from Irish fishery which lands a greater proportion of female Nephrops than either the French or Spanish fleet. The survey also shows higher proportions of females in the catches in the last few years.

There are no changes to other biological parameters for this stock and they are not relevant to the current trends based assessment.

## Surveys

The only fishery-independent source of data is from the Spanish Porcupine trawl survey. Further information on this survey is provided in the IBTS report (ICES, 2009) and in previous IBTS reports. Catchability of Nephrops in trawl surveys is typically an issue due to variable emergence patterns of Nephrops from their burrows (ICES, 2007). However, this stock (FU16) is found in deep water where animals are known to emerge mainly during the day. Survey hauls are only conducted during the day and the survey is scheduled for the same time each year, thus minimising variability due to emergence patterns. In addition, the Nephrops stock in this area is widely distributed and at relatively low densities over a large area, such that catchability is less variable than for those stocks in shallower water.

Problems with the trawl encountered in 2008 were rectified for 2009 and gear parameters and catch rates returned to more normal levels (WD 1). Distribution of Nephrops catches and biomass in Porcupine surveys between 2001 and 2009 are shown in Figure 7.6.6. There is evidence of an increase in abundance indices in 2009 particularly in one area of the ground (Figure 7.6.6). The stratified abundance estimate increased significantly in 2009 but it remains below levels observed in 2001 and 2002 (Figure 7.6.7). The biomass in 2009 shows a slight increase compared with 2008 but is also well below that observed at the start of the series.

The size structure of the catches in the survey shows two things: a much lower mean size than in the commercial fleets and an increasing trend in mean size for both sexes up to 2008 (Table 7.6.2, Figure 7.6.7). In 2009 there is large reduction of mean size in both sexes due to a recruiting year class with a modal length at around 27 mm . The proportion of larger sizes remains very low, $>3$ times lower than at the start of the series.

## Commercial cpue

The Nephrops fishery on the Porcupine Bank is both seasonal and opportunistic with increased targeting during periods of high Nephrops emergence and good weather.

Effort and lpue data are not standardised, and hence do not take into account vessel capacity, efficiency, seasonality or other factors that may bias perception of lpue and abundance trends over the longer term. These data are presented by country in Table 7.6.3 and Figure 7.6.9. Note: Irish and French effort is in hours Spanish effort is power adjusted and is reported in thousands of day*BHP/ 100 .

The effort index for the Spanish fleet (all gears) operating in Porcupine shows a steady decline from the 1970s until the early 1990s. Since then Spanish effort has declined more gradually. Nephrops lpue data for the Spanish fleet (all gears) shows a general declining trend until 2003. In 2004 and 2005 lpue increased rapidly, probably due to increased targeting of Nephrops, before declining again in the more recent years.

Fishing effort for French Nephrops vessels ${ }^{1}$ has fluctuated widely with peaks in the mid 1980s and through the late 1990s. Effort in 2009 was the lowest in the series. Lpue data for the French fleet in FU16 were high in the 1980s but declined with fluctuations to a series low in 2008.

Fishing effort data for the Irish otter trawl Nephrops directed fleet ${ }^{2}$. Increased rapidly over the period 2003-2007 before declining again in 2008 and 2009. Irish lpue has fluctuated but with a general declining trend.

### 7.6.3 Stock assessment

The assessment is based on multiple lines of evidence from several indicators. The available data includes commercial landings compositions for males and females from the main fleets. Catch rates and length distributions from the Spanish Porcupine Bank survey (2001-2009,) along with lpue and effort data for the main fleets.

[^15]
## Comparison with previous assessments

The assessment is based on similar indicators to those used in 2009. The additional data show a continued deterioration in stock status although there is new recruitment to the fishery and survey in 2009.

This year further information was provided from the Spanish Porcupine survey including spatial and size distributions of catches and gear parameters (WD 1). This fishery-independent information has proven increasingly important for this stock.

## State of the stock

The absolute stock size is uncertain but the stock is likely to be close or at the lowest levels observed based on stock indicators. Effort and landings trends indicate that fishing mortality has been high since the early 2000s. Fishery-independent survey information indicated that recruitment has been very weak or absent since 2004. However there is new recruitment to survey catches in 2009. This has been also been in the commercial catch data for males.

Landings per unit of effort (lpue) show a generally declining trend in most fleets over the time-series available and reached their lowest levels in the early 2000s. This probably reflects a decline in stock abundance. There was a substantial increase in landings and lpue in 2004 and 2005 indicating some signs of a stock increase, but since 2006 these indicators show a large decrease causing renewed concern about stock status. All the size distribution information shows a large increase in the size of Nephrops in this area. This is considered to be due to the combined effects of weak recruitment in recent years and the growth of a good year class that entered the commercial fishery in 2002. The combined effect of increased targeting and weak recruitment in recent years has resulted in a sudden deterioration in stock status.

Another important signal is the large change in sex ratio in the survey catches and fishery landings with female Nephrops accounting for a higher proportion since 2007. Such changes in sex ratio appear to be consistent with sperm limitation occurring at the population level. This occurs when the male component of the stock is reduced due to fishing leading to higher proportions of unfertilized females. Catchability of the unfertilized females increases as a consequence because they focus on feeding and growth rather than reproduction. Landings have declined by $17 \%$ between 2008 and 2009.

### 7.6.4 Short-term projections

There is no possibility to forecast catches in the short term using the available stock indicators.

### 7.6.5 MSY explorations

It has not been possible to carry out explorations of MSY targets for this stock but given the recent stock indicators the stock is probably exploited well above MSY levels.

### 7.6.6 Biological reference points

There are no reference points defined or agree for this stock.

### 7.6.7 Management plans

There is no management plan for this stock.

### 7.6.8 Uncertainties and bias in assessment and forecast

Discarding/high-grading practices for Nephrops fleets in this area are unknown and unquantified but all fleets show similar recent increases in mean size. All information points to poor recruitment and an increasing reliance of the fishery on larger individuals with a high female component.

### 7.6.9 Recommendation for next Benchmark

There needs to be improved sampling of catches for this stock. Sampling levels are currently low and several factors complicate sampling (see Section 7.6.2).

In the short term the survey may be the most appropriate method of monitoring stock status. The development of full analytical assessment would require better growth information and an improvement in sampling of catches. Spatially explicit landings and effort data, either by rectangle or at finer resolution by gear from all countries would also be useful.

Currently there are no plans to benchmark this stock before 2012.

### 7.6.10 Management considerations

Nephrops on the Porcupine Bank are fished in relatively deep waters over a widespread area where they occur at low abundance. Given the sedentary nature of Nephrops populations the closed area as introduced in 2010 may be an appropriate management tool to substantially reduce catches and allow the stock to recover the stock. The measure is expected to be quite effective at reducing fishing mortality provided that effort is not increase outside the time/area to compensate for the closure (Figure 7.6.10).

Productivity of deep-water Nephrops stocks is generally lower than that in shelf waters, though individual Nephrops grow to relatively large sizes and attain high market prices. Other deep-water Nephrops stocks off the Spanish and Portuguese coast have collapsed and have been subject to recovery measures for several years e.g. FU25, 26, 27 and 31. Recruitment in Nephrops populations in deep water may be more sporadic than for shelf stocks with strong larval retention mechanisms. This makes these stocks more vulnerable to over exploitation and potential recruitment failure as has been observed on the Porcupine Bank over the last decade.

### 7.6.11 References

ICES. 2007. Report of the Workshop on the use of UWTV surveys for determining abundance in Nephrops stocks throughout European waters (WKNEPHTV). ICES CM: 2007/ACFM: 14 Ref: LRC, PGCCDBS.

Table 7.6.1. - Porcupine Bank (FU 16): Landings (tonnes) by country, 1965-2009.

| Year | France | Rep. of Ireland | Spain | UK E \& W | UK Scotland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 514 |  |  |  |  | 514 |
| 1966 | 0 |  |  |  |  | 0 |
| 1967 | 441 |  |  |  |  | 441 |
| 1968 | 441 |  |  |  |  | 441 |
| 1969 | 609 |  |  |  |  | 609 |
| 1970 | 256 |  |  |  |  | 256 |
| 1971 | 500 |  | 1444 |  |  | 1944 |
| 1972 | 0 |  | 1738 |  |  | 1738 |
| 1973 | 811 |  | 2135 |  |  | 2946 |
| 1974 | 900 |  | 1894 |  |  | 2794 |
| 1975 | 0 |  | 2150 |  |  | 2150 |
| 1976 | 6 |  | 1321 |  |  | 1327 |
| 1977 | 0 |  | 1545 |  |  | 1545 |
| 1978 | 2 |  | 1742 |  |  | 1744 |
| 1979 | 14 |  | 2255 |  |  | 2269 |
| 1980 | 21 |  | 2904 |  |  | 2925 |
| 1981 | 66 |  | 3315 |  |  | 3381 |
| 1982 | 358 |  | 3931 |  |  | 4289 |
| 1983 | 615 |  | 2811 |  |  | 3426 |
| 1984 | 1067 |  | 2504 |  |  | 3571 |
| 1985 | 1181 |  | 2738 |  |  | 3919 |
| 1986 | 1060 |  | 1462 | 69 |  | 2591 |
| 1987 | 609 |  | 1677 | 213 |  | 2499 |
| 1988 | 600 |  | 1555 | 220 |  | 2375 |
| 1989 | 324 | 350 | 1417 | 24 |  | 2115 |
| 1990 | 336 | 169 | 1349 | 41 |  | 1895 |
| 1991 | 348 | 170 | 1021 | 101 |  | 1640 |
| 1992 | 665 | 311 | 822 | 217 |  | 2015 |
| 1993 | 799 | 206 | 752 | 100 |  | 1857 |
| 1994 | 1088 | 512 | 809 | 103 |  | 2512 |
| 1995 | 1234 | 971 | 579 | 152 |  | 2936 |
| 1996 | 1069 | 508 | 471 | 182 |  | 2230 |
| 1997 | 1028 | 653 | 473 | 255 |  | 2409 |
| 1998 | 879 | 598 | 405 | 273 |  | 2155 |
| 1999 | 1047 | 609 | 448 | 185 |  | 2290 |
| 2000 | 351 | 227 | 213 | 120 |  | 910 |
| 2001 | 425 | 369 | 270 | 158 |  | 1222 |
| 2002 | 369 | 543 | 276 | 139 |  | 1327 |
| 2003 | 131 | 307 | 333 | 108 | 29 | 908 |
| 2004 | 289 | 494 | 588 | 126 | 28 | 1526 |
| 2005 | 397 | 754 | 799 | 208 | 156 | 2315 |
| 2006 | 462 | 731 | 571 | 201 | 155 | 2120 |
| 2007 | 302 | 1060 | 496 | 146 | 183 | 2186 |
| 2008 | 26 | 562 | 234 | 41 | 138 | 1000 |
| 2009 | 4 | 356 | 294 | 12 | 159 | 825 |

Table 7.6.2 - Porcupine Bank (FU 16): Mean sizes (mm CL) of male and female Nephrops in Spanish, French and Irish landings and the Spanish Porcupine Groundfish survey 1981-2009

|  | Spain |  | Rep. Of Ireland |  | France |  | Porcupine Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings |  | Landings |  | Landings |  | Catch |  |
|  | Males | Females | Males | Females | Males | Females | Males | Females |
| 1981 | 39.9 | 34.5 | - | - | - | - | - | - |
| 1982 | 40.9 | 34.8 | - | - | - | - | - | - |
| 1983 | 40.8 | 34.0 | - | - | - | - | - | - |
| 1984 | 39.7 | 33.1 | - | - | - | - | - | - |
| 1985 | 38.7 | 33.5 | - | - | - | - | - | - |
| 1986 | 40.7 | 36.4 | - | - | - | - | - | - |
| 1987 | 39.3 | 35.0 | - | - | - | - | - | - |
| 1988 | 40.7 | 38.3 | - | - | - | - | - | - |
| 1989 | 40.5 | 36.8 | - | - | - | - | - | - |
| 1990 | 41.0 | 36.1 | - | - | - | - | - | - |
| 1991 | 39.4 | 34.5 | - | - | - | - | - | - |
| 1992 | 39.2 | 34.1 | - | - | - | - | - | - |
| 1993 | 41.6 | 36.1 | - | - | - | - | - | - |
| 1994 | 40.8 | 36.5 | - | - | - | - | - | - |
| 1995 | 41.3 | 36.6 | 40.7 | 36.5 | 43.2 | 38.3 | - | - |
| 1996 | 41.6 | 35.1 | 34.6 | 35.3 | 41.7 | 38.9 | - | - |
| 1997 | 39.7 | 34.8 | 35.9 | 34.5 | 41.9 | 38.4 | - | - |
| 1998 | 41.1 | 34.6 | 37.2 | 35.6 | 41.9 | 38.4 | - | - |
| 1999 | 41.5 | 35.7 | 36.6 | 33.7 | 43.1 | 39.1 | - | - |
| 2000 | 41.1 | 34.8 | na | na | 45.3 | 40.5 | - | - |
| 2001 | 41.1 | 36.3 | 37.8 | 35.4 | 45.4 | 39.4 | 35.5 | 28.4 |
| 2002 | 39.7 | 35.3 | 36.1 | 38.5 | 45.3 | 40.3 | 37.0 | 31.2 |
| 2003 | 41.4 | 37.8 | 44.5 | 36.2 | 46.2 | 38.9 | 39.2 | 31.4 |
| 2004 | 43.5 | 38.5 | 43.5 | 35.7 | 46.4 | 41.5 | 39.4 | 30.0 |
| 2005 | 43.4 | 38.1 | 46.9 | 40.6 | 45.9 | 41.0 | 44.6 | 33.3 |
| 2006 | 43.9 | 38.0 | na | na | 48.9 | 41.4 | 43.6 | 34.5 |
| 2007 | 43.7 | 41.0 | na | na | 48.3 | 43.8 | 45.4 | 37.4 |
| 2008 | 51.0 | 40.6 | 43.3 | 37.5 | na | na | 48.0 | 38.2 |
| 2009 | 43.0 | 42.7 | 44.1 | 40.1 | na | na | 32.2 | 28.3 |

Table 7.6.3. - Nephrops Porcupine Bank (FU 16)
Landings and effort for the various different fleets exploiting the stock 1971-2009

| Year | Spanish fleet |  |  | French Nep fleet ${ }^{1}$ |  |  | Irish Nep Fleet ${ }^{2}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Landings | Effort |  | Landings | Effort | LPUE (>10\%) | Landings | Effort | LPUE |
|  | Tonnes | $\begin{gathered} \text { day*BHP/100 } \\ (\times 1000) \end{gathered}$ | $\begin{gathered} \hline \text { T/day * } \\ \text { BHP/100 } \end{gathered}$ | Tonnes | ('000's Hrs) | (kg/hr) | Tonnes | ('000's Hrs) | (kg/hr) |
| 1971 | 1444 | 159 | 9 |  |  |  |  |  |  |
| 1972 | 1738 | 188 | 9 |  |  |  |  |  |  |
| 1973 | 2135 | 181 | 12 |  |  |  |  |  |  |
| 1974 | 1894 | 192 | 10 |  |  |  |  |  |  |
| 1975 | 2150 | 229 | 9 |  |  |  |  |  |  |
| 1976 | 1321 | 187 | 7 |  |  |  |  |  |  |
| 1977 | 1545 | 196 | 8 |  |  |  |  |  |  |
| 1978 | 1742 | 166 | 11 |  |  |  |  |  |  |
| 1979 | 2255 | 157 | 14 |  |  |  |  |  |  |
| 1980 | 2904 | 163 | 18 |  |  |  |  |  |  |
| 1981 | 3315 | 143 | 23 |  |  |  |  |  |  |
| 1982 | 3931 | 138 | 29 |  |  |  |  |  |  |
| 1983 | 2811 | 108 | 26 | 615 | 18 | 35 |  |  |  |
| 1984 | 2504 | 114 | 22 | 1067 | 30 | 35 |  |  |  |
| 1985 | 2738 | 115 | 24 | 1181 | 33 | 36 |  |  |  |
| 1986 | 1462 | 95 | 15 | 1060 | 28 | 38 |  |  |  |
| 1987 | 1677 | 105 | 16 | 609 | 24 | 26 |  |  |  |
| 1988 | 1555 | 109 | 14 | 600 | 22 | 27 |  |  |  |
| 1989 | 1417 | 105 | 14 | 324 | 14 | 23 |  |  |  |
| 1990 | 1349 | 96 | 14 | 336 | 15 | 23 |  |  |  |
| 1991 | 1021 | 85 | 12 | 348 | 19 | 18 |  |  |  |
| 1992 | 822 | 59 | 14 | 665 | 32 | 21 |  |  |  |
| 1993 | 752 | 49 | 15 | 799 | 36 | 22 | 206 |  |  |
| 1994 | 809 | 50 | 16 | 1088 | 38 | 28 | 512 |  |  |
| 1995 | 579 | 48 | 12 | 1234 | 42 | 30 | 971 | 15 | 41 |
| 1996 | 471 | 43 | 11 | 1069 | 41 | 26 | 508 | 8 | 42 |
| 1997 | 473 | 42 | 11 | 1028 | 41 | 25 | 653 | 11 | 35 |
| 1998 | 405 | 43 | 10 | 879 | 40 | 22 | 598 | 10 | 42 |
| 1999 | 448 | 37 | 12 | 889 | 43 | 21 | 609 | 9 | 35 |
| 2000 | 213 | 30 | 7 | 313 | 23 | 16 | 227 | 2 | 31 |
| 2001 | 270 | 29 | 9 | 366 | 24 | 17 | 369 | 8 | 30 |
| 2002 | 276 | 31 | 9 | 324 | 18 | 22 | 543 | 10 | 38 |
| 2003 | 333 | 38 | 9 | 130 | 7 | 19 | 296 | 7 | 26 |
| 2004 | 588 | 32 | 18 | 232 | 9 | 25 | 494 | 16 | 21 |
| 2005 | 799 | 30 | 27 | 380 | 15 | 26 | 628 | 24 | 30 |
| 2006 | 571 | 39 | 15 | 446 | 22 | 21 | 683 | 28 | 25 |
| 2007 | 496 | 35 | 14 | 297 | 17 | 20 | 977 | 36 | 27 |
| 2008 | 234 | 24 | 10 | 25 | 4 | 7 | 534 | 20 | 26 |
| 2009 | 294 | 26 | 11 | na | na | na | 327 | 12 | 27 |
| ${ }^{1}=$ Vessels where $<10 \%$ of landed value was Nephrops; ${ }^{2}=$ Vessels where $30 \%$ of the landed weight was Nephrops |  |  |  |  |  |  |  |  |  |



Figure 7.6.1. Nephrops in FU16 (Porcupine Bank). Landings in tonnes by country.


Figure 7.6.2. Nephrops in FU16 (Porcupine Bank). Landings mean sizes by sex and country and mean size in the catch for the Porcupine survey.

## Length frequencies for Landings:

Nephrops in FU16


Figure 7.6.3. Nephrops in FU16 (Porcupine Bank). Female and male landings length distributions.


Figure 7.6.4. Nephrops in FU16 (Porcupine Bank). Trends in various indicators from male length frequency data.


Figure 7.6.5. Nephrops in FU16 (Porcupine Bank). Sex ratio of landings and survey catches.

Nephrops norvegicas


Figure 7.6.6. Nephrops in FU16 (Porcupine Bank). Distribution of Nephrops norvegicus catches in Porcupine surveys between 2001 and 2009.


Figure 7.6.7. Nephrops in FU16 (Porcupine Bank). Changes in Nephrops norvegicus biomass and number stratified indices during Porcupine Survey time-series (2001-2009). Boxes mark parametric standard error of the stratified abundance index. Lines mark bootstrap confidence intervals ( $\alpha=0.80$, bootstrap iterations $=1000$ ).

## Length frequencies for Landings:

Nephrops in FU16


Figure 7.6.8. Nephrops in FU16 (Porcupine Bank). Female and male Porcupine Survey length distributions.

Figure 7.6.9 Nephrops in FU 16 (Porcupine Bank) Effort and LPUE trends for fieets
(') The Spanish effort index is based on a combination of hous at sea and average engive power. Irish and French effort and LPUE is unstandardised.




Figure 7.6.10. Nephrops in FU16 (Porcupine Bank). The area on the Porcupine Bank to be closed seasonally to Nephrops fishing in 2010 shown as a dotted black line overlaid on the distribution of recent (2006-2006) Irish fishing effort directed towards Nephrops.

### 7.7 Nephrops in the Celtic Sea, FU20-22

ICES description VIIfgh
Functional Units Celtic Sea, VIIfgh (FU20-22)
Type of assessment in 2010
Update assessment.
The assessment in 2010 is based on an examination of trends in fishery indicators. Main changes in the assessment methodology compared to last year: discard derivation investigated for French trawlers (WGSSDS 2006-2008; WGCSE 2009; Stock Annex) is temporarily delayed because of the revision for DLFs from French landings including tailed individuals. Unknown lpue status in 2009 due to the lack of reliable information from official French statistics of Fisheries Direction. UWTV survey data is also available for the "Smalls" component of FU20-22 but this is not representative of the whole stock area.

ICES advice applicable to 2009 and 2010
Exploitation boundaries in relation to precautionary considerations
"The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to 2007. This corresponds to landings of no more than 5300 tonnes for the Celtic Sea stock (FU20-22)."

### 7.7.1 General

## Stock description and management units

The Celtic Sea Nephrops stock (FU20-22) is included in the whole ICES Area VII as Irish Sea East [FU14], Irish Sea West FU15], Porcupine Bank [FU16], Aran Islands [FU17], northwest Irish Coast [FU18], southeast and southwest Irish Coast [FU19]. The TAC is set for Subarea VII which does not correspond to the stock area.

There is no evidence that the whole exploited area belongs to the same stock or that there are several patches linked in meta-population sense.


The TAC for Subarea VII is bounded by the red line. The FUs with the TAC area are shaded.

## Management applicable in 2008 and 2009

Currently the TAC is set for Subarea VII .The 2010 TAC is $22432 \mathrm{t}, 9 \%$ less than the 2009 TAC. This TAC includes many Nephrops stocks and this may allow unrestricted catches for stocks under excessive fishing pressure where catches should be limited.

The MLS implemented by EC is set at 25 mm CL i.e. 8.5 cm total length and this regulation is applied by the Irish and UK fleets whereas a more restrictive regulation adopted by the French Producers' Organisations ( 35 mm CL i.e. 11.5 cm total length) is applied by the French trawlers.

In application of the Council Regulation (EC) $\mathrm{N}^{\circ}$ 1459/1999, June 24th, 1999, modifying the regulation (EC) $\mathrm{N}^{\circ} 850 / 98$ of the Council for the conservation of fishery resources through technical measures for the protection of juveniles, the French minimum mesh size of codend was set at 100 mm in January 2000 whereas the Irish mesh size was maintained at 80 mm .

## TAC in 2009

| Species: Norway lobster Neplopops narvegkus |  | Zone | $\begin{aligned} & \text { VII } \\ & \text { NEPY(07) } \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Spain | 1479 |  |  |
| France | 5994 |  |  |
| Ireland | 9091 |  |  |
| United Kingdom | 8086 |  |  |
| EC | 24650 |  |  |
| TAC | 24650 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EQ) No 847/96 applies. |

TAC in 2010

| Species:Norway lobster <br> Nephrops narvegicus | Zone:VII <br> (NEP/07) |  |
| :--- | :---: | :--- | :--- |
| Spain | 1346 |  |
| France | 5455 |  |
| Ireland | 8273 |  |
| United Kingdom | 7358 |  |
| EU | 22432 | Analytical TAC |
| TAC | 22432 |  |

## Ecosystem aspects

This section is detailed in Stock Annex.

## Fishery description

This section is detailed in Stock Annex.

## Fishery in 2009

France and Ireland are the main countries involved in the FU20-22 Nephrops fishery.
In 200974 French trawlers landed Nephrops from FU20-22 (88 in 2008). Of these, 43 exceeded landings of 10 t representing more than $98 \%$ of French landings. Five vessels accounted for more than $20 \%$ of the total quantity harvested by France where these vessels landings were between 78 t and 103 t . In 2009, 79 Irish vessels reported landings from FU20-22 (99 in 2008). Of these, 54 vessels (67 in 2008) reported landings in excess of 10 t accounting for $95 \%$ of the total landings.

A decommissioning programme was in operation in Ireland during 2007 and 2008. Twelve vessels active in the FU 20 fishery were decommissioned. These vessels accounted for approximately $18 \%$ of the landings in the 2007-2008 period.

In 2009 increasing fuel prices and poor market value for Nephrops impacted in the activity of the fleets.

## Uptake of quotas

There is no specific TAC or quota for the FU20-22 Nephrops, thus the question should be examined for the whole Subarea VII. For the two main fleets operating in the Celtic Sea, the total harvested quantities in Area VII remained below the allowed quotas. In 2008, 5994 t were allocated to France whereas actual French landings 2008 were 2420 t almost exclusively i.e. $97 \%$ coming from the Celtic Sea. In 2008, 9091 t were allocated to the Republic of Ireland and 9053 t were landed ( $38 \%$ from the Celtic Sea). For 2009, the ICES recommended for the FU20-22 Nephrops to not exceed 5300 t of landings coming: however, the total harvested quantity was 5359 t . This value may be revised when current information from the official French statistics on landings is modified.

### 7.7.2 Data

An overview of the data provided and used by the WG is shown in Table 2.1.

## Landings

At WGSSDS 2008, the French landings were revised sharply downwards for 1999 ( 2036 t against 2745 t previously used). This revision was performed by including information of the new database "Harmonie" of statistics. The change involves notably reported fishing effort of trawlers from the FU20-22 stock (Celtic Sea) to the FU16 (Porcupine Bank) during the 2nd quarter of this year. In 2009, new revisions of French and of Irish statistics were carried out by the WGCSE (see Table below).

| Country | Year | Previous landings (t) | Revised landings (WGCSE 2009) (t) |
| :---: | :---: | :---: | :---: |
| France | 1999 | 2036 | 2078 |
|  | 2000 | 2782 | 2848 |
|  | 2001 | 2532 | 2626 |
|  | 2002 | 3134 | 3154 |
|  | 2003 | 3510 | 3595 |
|  | 2004 | 2511 | 2605 |
|  | 2005 | 2490 | 2502 |
|  | 2006 | 2397 | 2368 |
|  | 2007 | 2082 | 2033 |
| Ireland | 2006 | 1877 | 1864 |
|  | 2007 | 3226 | 3213 |

The new official French statistics were generally changed upwards apart from years 2006 and 2007 whilst a proportion of the fishing effort for French trawlers was reallocated to the FU19 (SE and SW Irish coast).In 2010, Irish landings for 2008 were slightly modified ( 3411 t against 3422 t ).

Landings are reported mainly by France and the Republic of Ireland (Figure 7.7.1; Table 7.7.1). The contribution of French landings has gradually decreased from 80$90 \%$ at the end of 1980 s to $50-60 \%$ at the beginning of 2000 s. Then, since 2007 , French landings declined to less than $40 \%$ of the total reported quantities (Table 7.7.1). The overall fishing profile remains typically seasonal (Table 7.7.2) with the majority of landings coming from the 2 nd and 3rd quarters.

French landings increased notably by $16 \%$ from 2033 t in 2007 to 2348 t in 2008. In 2009 landings decreased by $8 \%$ to 2156 t which is the third lowest value in the French time-series.

Irish landings in 2008 ( 3411 t ) were the highest reported for this nation's time-series. In 2009, landings declined by around $17 \%$ to 2844 t (third highest in the series).

Table 7.7.2 shows the landings by quarter for France and Ireland and there is a seasonal trend exhibited by both countries mainly in quarters 2 and 3.

## Commercial Ipue

Effort data is available from 1983 to 2008 for the French Nephrops fleet (Table 7.7.17; Figure 7.7.9). No 2009 data is available because the new registration system of official French statistics is not yet validated.

French effort has fluctuated over the series with a decreasing trend since 2004 to the lowest observed in 2008 the available time-series. The decrease of the French fishing effort was caused by the reduction of the number of trips by vessel whereas the total number of vessels remained almost stable. Lpue for French trawlers increased between 2007 and 2008 ( $+22 \%$ : $22.6 \mathrm{~kg} / \mathrm{h}$ in 2008 against $18.5 \mathrm{~kg} / \mathrm{h}$ for 2007). In 2009, because of lack of reliable official French statistics as explained above, lpue indices are unavailable.

Effort data for this FU is available from 1995 for the Irish otter trawl Nephrops directed fleet. These data have not been standardised to take into account vessel or efficiency changes during the time period. Irish effort has fluctuated over the series with an increasing trend since 2004 to 2008 and declined somewhat in 2009 (Table 7.7.17; Figure 7.7.9). The increase of the Irish fishing effort involves either in the number of fishing vessels ( 95 and 99 Irish trawlers were respectively listed in 2007 and 2008 compared to 80 for 2006) or in the number of trips by vessel. Lpue has been fluctuating around an average of $45 \mathrm{~kg} / \mathrm{h}$ with an increase to the highest ( $54 \mathrm{~kg} / \mathrm{h}$ ) in the time-series in 2008. A decrease occurred in 2009 to $48 \mathrm{~kg} / \mathrm{h}$.

## Biological

Length-frequency distribution information by country (France and Ireland) is given in the Stock Annex. All data are presented in and Tables 7.7.3 to 7.7.10, 7.7.11, 7.7.12 and 7.7.13a. The Table 7.7.14 provides information on mean size of landings by year and country.

The LFD data reveal significant differences between the two countries. The two ogives of selectivity through meshes are different. The evolution of the French landings had shown a substantial increase of mean sizes since the beginning of 2000s (this coincides with mesh regulations cited in the Stock Annex), but a significant decrease of the mean size occurred in 2006 ( 41.0 mm CL for both sexes combined against 42.8 mm CL in 2005). The same trend was observed for Irish landings ( 29.2 mm CL against 31.1 mm CL in 2005). In 2007, a new decrease occurs for both countries ( 40.4 mm CL for France, 28.4 mm for Ireland).

The WGCSE 2009 and 2010 pointed out a significantly increasing proportion of tailed individuals present in French landings (Figure 7.7.3) whereas this proportion was already high for Irish trawlers. For years 2005-2008, tailed Nephrops were comprised between 11 to $20 \%$ of the French landings whereas this component of the landings was less than $5 \%$ until the beginning of 2000s. In 2009, the tailed Nephrops component accounted for $17 \%$ of French landings. Industry explained this recent change due to the economic difficulties of increasing fuel prices. Tailed individuals are intended to
compensate this loss for the crew participation at the total investment by trip. As the European MLS for FU20-22 Nephrops is fixed at 8.5 cm of total length ( 25 mm CL) and the MLS retained by the French Producers' Organizations is equal to $11.5 \mathrm{~cm}(35 \mathrm{~mm}$ $\mathrm{CL})$, it was expected that tailed individuals should be comprised between these two sizes.

By the end of 2007, tailed Nephrops could not be sampled at auction and, as the sampling onboard remains difficult to apply routinely due to long trip duration by the French trawlers, the problem was partially tackled by apportioning tailed individuals to the smallest category of landings at auction. Since the end of 2007, new biometric relationships established during the EVHOE survey have been used (Stock Annex): this allows fitting CL vs. 2nd abdominal segment of tail by sex.

The DLF of French landings for 2008 and 2009 were estimated by two ways: one using the extrapolations from tails to CL, the other apportioning tails to the small category as for previous years. The resulting difference appears relevant (Table 7.7.12; Figures 7.7.4 and 7.7.5): in 2008, 46 million Nephrops were provided by the previous method whereas 58 million were estimated by including tails (+28\%). Almost $30 \%$ of landed individuals were below the French Producers' Organization MLS, but no Nephrops was undersized compared with European MLS. Moreover, the sex ratio seems to be affected by the tailing practice: $13 \%$ of Nephrops ( 7.4 million) were females although this percentage would be $7 \%$ ( 3.2 million) under the previous method. In 2009, the method including tails and the previous one give respectively 48 and 39 million Nephrops (+21\%); moreover, 19\% of individuals are smaller than the French Producers' Organization MLS and the sex ratios are respectively $10 \%$ and $3 \%$. As indicated in Table 7.7.14 the mean size of French landings for 2008 and 2009 decreases at around $2.5-5.5 \mathrm{~mm}$ CL by sex when tails are involved by sampling. However, the mean CL for 2008 remains larger than the Irish one.

The size composition is overestimated when raised to the composition of entire individuals and, therefore, the total number of landed Nephrops is underestimated.

## Discards

The increasing practice of tailing Nephrops for the French trawlers may affect the total discard rate of this fleet. Hence, method for discard derivation applied since 2006 on DLF French dataset for years with no sampling onboard is not currently used for the assessment.

## Sampling

The available dataset is detailed in the Stock Annex. Additional French dataset was also acquired in 2005, but it involves in only two quarters (Q3 and Q4; Stock Annex). Data sampled in 2009 ( 14 trips, 199 hauls in three quarters) cannot yet be routinely integrated in the assessment. As for landings, the Irish discard sampling began in 2002. Thus, there is no common dataset on discards between French and Irish fleets (lack of information of the Irish sampling program for $2005-\mathrm{Q} 3,2005-\mathrm{Q} 4,2006-\mathrm{Q} 4$, 2007-Q2). Available information on complete yearly sets (1997-FR, 2003-IRL, 2008IRL, 2009-IRL) is given by Figures 7.7.6, 7.7.7 and Table 7.7.15. Tables 7.7.13b,c,d provide discard estimates, total catches and removals for Irish trawlers (using mortality rate of discards equal to $75 \%$ : Charuau et al, 1982).

The notable contrast between the retained proportions onboard and the spatial heterogeneity of the exploited area prevents direct comparisons of the main fleets. It is not yet possible to estimate if the inter-fleet variability of the discard rate is larger than the inter-annual one.

Changes in discard rate is a consequence of the strength of recruitments, increase in the MLS (which tends to increase the discards) and the gear selectivity. Other practices as stated above (tailing individuals) may affect discard rate. The relative contribution of each of these four factors remains unknown.

## Back-calculation

As for the main Nephrops stocks, the lack of estimation of discards hampers quantitative analysis of recruitment indices, therefore, possibilities of back-calculation for discards were investigated. For a long period, a "proportional derivation" of discards was processed on the FU20-22 Nephrops by WGNEPH, but was considered as unreliable because it induces lack of contrast in inter-annual variations of recruitment (see reports of WGSSDS 2005-2008; WGCSE 2009). An alternative probabilistic approach developed since 2006 on other Nephrops stocks (VIIIab; Bay of Biscay; FU23-24) was also applied to the FU20-22. The main concepts of the back-calculation are detailed in Stock Annex.

The increasing proportion of tails probably modifies discard practices and the WGCSE 2010 decided not to develop the back-calculation approach as this stock is not benchmarked.

## Surveys

## UWTV Survey

The UWTV survey was developed in 2006 by Ireland and has become the main source of fishery independent and new information on this stock. The methods employed during the Celtic Sea UWTV surveys have recently been discussed and well documented by WKNEPHTV, WKNEPHBID, WKNEPH and SGNEPS (ICES, 2007, 2008, 2009a and 2009b).This survey indicates that burrow density in 2009 for the "Smalls" ground decreased slightly $\sim 2 \%$ from that observed in 2008, but has remained stable (Working Document 9).

Figure 7.7.8 points to a possible relationship between the 2006 geostatistical abundance estimate which is the highest level observed in the series and the strong lpue values obtained by commercial vessels in the same area in 2007. However, a more detailed investigation is required to examine the relationship between burrow abundance and lpue. The survey area is not representative of the whole stock area and there are many discrete Nephrops grounds with FU20-22. Table 7.7.16 gives the landings by statistical rectangle and shows that the Irish and French fisheries exploit different grounds. By the end of 1990s, more than $40 \%$ of French landings were reported from the "Smalls" area compared to less than $10 \%$ at the end of 2000s, whereas $2 / 3$ of the total Irish landings are reported from the "Smalls" (ICES statistical rectangle 31E3) WGCSE propose that the various stock parameters and reference points required to develop catch options for FU20-22 "Smalls" component will be explored during the summer and made available to apply to the new survey data when a new data point will be available after July 2010.

In FU20-22, the French groundfish survey EVHOE while not focusing on Nephrops per $s e$, it does provide some indication of the length distributions and the strength of recruitment (Stock Annex). The Irish groundfish survey has been carried out since 2003 giving some information on the length compositions of Nephrops catches. The UK bottom trawl survey occurred on the same area between 1984 and 2004 (see WGSSDS 2006), however, only two sampling stations were surveyed within FU20-22 area. Further information on these surveys is provided in the IBTS report (ICES, 2009) and in previous IBTS reports.

In 2008, an experiment intending to update the maturity ogive for females was carried out during the EVHOE survey. The sampling plan of this survey is not designed for Nephrops (late period in the year i.e. November/December affecting female abundance because of burrowing) and also the sensitive period for Nephrops maturity occurs during the 2 nd and 3rd quarters. These data should be collected during a more adequate season; however, data sampled owing to commercial trips may be biased (selectivity of mesh size 100 mm ), furthermore the long duration of French trips is an obstacle for this experiment.

## Other relevant data

French partnership of the fishing industry underlined that the increase of lpue series since the end of 1990s may be caused by the change of the global fishing efficiency of the fleet because some old vessels were replaced by more recent ones. Fishing power analysis including spatial distribution will be undertaken on a set of French Nephrops trawlers remaining in the fishery for a long period (e.g. 1999-2008; 40 vessels) combining information involving in other substantial species targeted in the Celtic Sea (cod). Furthermore, the problem of the actual size composition of tailed individuals in landings was also debated with Producers' Organisations. The possibility of European regulation such as a numerous clauses licence system was also debated. Moreover, taking into account the current difficulty to collect information during French commercial trips, the perspective of self-sampling applied on discarded fraction of catches was also discussed.

### 7.7.3 Historical stock development

For a long period, the FU20-22 Nephrops stock was analytically investigated by XSA. However, the Nephrops ageing cannot be performed routinely. The L2AGE slicing program is usually applied on Nephrops stocks and allocates length classes into age groups by assuming von Bertalanffy model of individual growth. This slicing can be applied to length distributions by sex. All parameters, $\mathrm{L} \infty$ and K by sex, calculated mean sizes by age for each sex, natural mortality and maturity by sex (assumed to be knife-edged for males and s-shaped for females) and combined are given in Stock Annex.

The slicing process converting size-frequencies to age-compositions at the aim of performing XSA is often disapproved because it may induce lack of contrast between years (input set of common parameters for individual growth). Moreover, the von Bertalanffy's equation is often invalidated for crustaceans. As it would not be reasonable to expect that methods of direct age determination for Nephrops will be routinely available in the foreseeable future, alternative methods as CSA have to be investigated. The main current disadvantage of CSA is linked to the recruitment indices required: as the independent UWTV survey cannot yet provide consistent dataset on young year classes, the recruitment indices can be given only by annual discard indicators. Thus, no CSA investigation can be envisaged before providing discards estimators by reliable indices for both main fleets and for the complete time-series.

## Comparison with previous assessments

Only comparisons based on global indicators for the stock can be carried out. Even if there is no possibility for catch-at-age analysis regarding absolute levels of abundance of Nephrops in FU20-22, there is usually significant information on the relative stock state.

The current abundance indices applied for both main fleets for FU20-22 involve in commercial data (thresholds of $10 \%$ and $30 \%$ i.e. percentage of total landings composed from Nephrops respectively applied on French and Irish trawlers as tuning fleets).

For the WGCSE 2010, the information is severely hampered due to the lack of 2009 official French data on fishing effort and lpue whereas information on landings based on the sales at auction seems to be more satisfactory.

In 2007, the lpue has increased substantially for the Irish fleet and slightly for the French fleet. In 2008, both lpue changed upwards mainly the French one. The EVHOE also shows a strong increase in catches for 2007 and 2008. In 2009, Irish indices declined and there is no possibility for calculation of French indices as explained above.

The French trawlers' lpue and cpue series both have indicated a rise in stock abundance since the early 2000s. However, it should be important to investigate whether the sharp Irish lpue variations correspond to actual signals for the stock or to other factors linked to fleet capacity. It is noticeable that the French groundfish survey EVHOE while no focusing on Nephrops had provided in 2007 the highest indices for this species since the beginning of the survey 10 years ago. Trenkel and Rochet (2003) examining indicators in the French EVHOE Celtic Sea survey suggest that Nephrops population is increasing during 2000s.

Until 2005, the mean size of landings had also increased except for 2001 when the smaller size composition suggests a stronger recruitment entry in the fishery. Nevertheless, in 2006 and 2007, mean sizes in landings for both fleets decreased. This point combined to the former UK survey on this area (suggesting a slight trend of decrease of mean sizes for some sampling reference stations: see WGSSDS 2006) could be induced either by stronger recruitment abundance than previously or by over-fishing.

From the end of 2007 onwards, the revised DLFs taking into account tailed individuals in French sampling changed estimates of mean sizes: WGCSE 2010 performed recalculation of the mean sizes and sex ratios over the period since 1999. On 28314 t of official French landings over the period 1999-2009, the actual number of landed Nephrops is estimated equal to 591 million whereas the previous sampling process provided 515 million ( $+15 \%$ ). It is not currently possible to estimate whether the additional removals up to $+15 \%$ contradict the current advice on this stock mostly if the additional landings are compensated by less discards.

As no analytical XSA run was performed, abundance of recent recruiting year classes can be examined only by comparison of independent indicators such as discarded individuals estimated by the logistic derivation method and some surveys indices. As detailed in the Stock Annex, independent sources of information (EVHOE survey's indices, logistically derived discards for no sampled years) agree that some recent recruiting classes (mainly 2001 and probably 2002 and 2003) should be of a good level whereas it is still impossible to indicate the actual state of the more recent year classes.

## State of the stock

The state of the stock is unknown.

### 7.7.4 Short-term projections

No short-term projection was performed for this stock. For other Nephrops stocks in VII with UWTV surveys these have been used as the basis for catch options in 2011.

This should be also possible for the Smalls component of FU20-22 (32E3, 31E3 and 31E4) which represents around $50 \%$ of the total landings (average on years 20032008). A new survey estimate will be available in late summer 2010. WGCSE propose that the various stock parameters and reference points required to develop catch options for FU20-22 will be explored during the summer and made available to apply to the new survey data.

### 7.7.5 MSY explorations

No MSY explorations were carried out at WGCSE but these explorations will be carried out for the Smalls component of the catches during the summer of 2010 in time for the 2011 advice.

### 7.7.6 Biological reference points

There are no biological reference points for FU20-22 Nephrops stock.

### 7.7.7 Management plans

No specific management plan exists for this stock.

### 7.7.8 Uncertainties and bias in assessment and forecast

The revision of French landings, fishing effort and lpue over the recent years, underlines the heterogeneous composition of the standard pool of vessels (e.g. it could be divided into two separate fleets i.e. the one able to switch between different stocks of the Subarea VII and the other composed of less efficient trawlers limited to the Celtic Sea). Currently, misreporting does not seem to be a problem for the stock.

## Exploitation pattern and spatial variability

The French and Irish time-series remain different and were provided by applying different exploitation pattern on different areas. As pointed out by the Table 7.7.16, French and Irish trawlers cover different areas and have presented contrasting features over the last decade. French fleet moved gradually from the "Smalls" Ground (32E3, 31E3 and 31E4) to the "Labadie" (30E2): at the end of 1990s, more than $40 \%$ of French landings were reported from the "Smalls" area whereas the contribution of this rectangle became minor less than $10 \%$ at the end of 2000s. Irish vessels fishing occupied the "Smalls" ground (current production of 31 E 3 is around $2 / 3$ of the total Irish landings and around $50 \%$ of the total landings from FU20-22).

## Heterogeneity of DLFs for landings and discards

The problem of high variability of landing samples between trips still remains (higher coefficients of variation at auction because of higher heterogeneity of the fished area and of long duration of trips i.e. 12-15 days and, therefore, less availability of samples at auction). Hence, high CV of numbers at sizes (20-30\%) are usual. In any case, commercial samples can be extended by including the commercial part sampled onboard during the DCF plan.

The sampling of tailed individuals in French landings provides valuable information, but underlines the necessity to re-calculate the actual size-composition of discarded individuals under the revised DLFs for landings.

While the selectivity parameters are not significantly improved for Nephrops trawlers, it appears appropriate to continue the Irish discard plan and to conduct a French one
on a yearly basis. For French trawlers, it should be suitable to investigate possibilities of reliable self-sampling onboard. It should be interesting to examine the part of decrease of the French discard rate since the early 2000s due to the selectivity improvement from that related to some weak recruiting classes (however, size-composition of landings for 2006 and 2007 may suggest a positive signal for recruitment). Moreover, if the individual growth of this species is faster during the latter period of the compiled time-series, there would be decline of the discarded amounts with no possibility to investigate the actual recruitment level.

### 7.7.9 Recommendation for next Benchmark

Many quantitative explorations attempted in recent years for the FU20-22 Nephrops stock (e.g. sampling onboard, maturity ogive, discard derivation) were handicapped by the overall spatial heterogeneity, by the divergence of exploitation pattern for the two main fleets.As the stock state seems to be stable, fishery indicators only provide provisionally adequate information. The following issues need to be addressed at the benchmarking of this stock.

## Biological sampling

## Auction

As the French sampling of tailed Nephrops on landings at auction has recently been standardized, updated information for DLF and sex ratio was provided in 2010 and should be benchmarked.

## Onboard

The Irish plan of sampling onboard under DCF will continue to provide information on discarded amounts and DLF. For the French trawlers, self-sampling onboard may be more realistic than in the past (concentration of a huge proportion of total landings from a small number of vessels; see above §7.7.1). Difficulties of sampling onboard when long trip duration should be addressed.

## Maturity

Re-estimation of maturity parameters requires a specifically designed experiment which should be commonly organized by France and Ireland under DCF.

## Back-calculation for missing biological data

## Tails

The modification of DLF for tailed individuals was extended on the overall period since the tailed fraction became significant by applying probabilistic concepts combined with s-shaped quarterly curves of tailing Nephrops vs. size.

## Discards

After re-calculating DLF for French landings on recent years, by an analogous way as already performed by WGSSDS 2006-2008, DLF of discards for French trawlers should be carried out for the whole time-series integrating the change of relative selectivity for trawls in 2000 ( 100 mm replacing 80 mm ).

## Dataset on DLF of Irish landings before 2002

For the years 1995-2002, available series on Irish landings on quarterly basis was not associated to samples on DLF. In spite of spatial variability affecting size composition by fleet, the possibility to extrapolate French DLF for this period has to be investigated: before 2000, the same selectivity parameters for trawls should be used (the difference involved in MLS; §7.7.1).

## Surveys

## UWTV Irish survey

The UWTV Irish survey initiated in 2006 this could form the basis for catch options and management advice for the the Smalls component of the stock using methodologies outlined in WKNEPH 2009. This will be carried out later in 2010 outside the scope of a full benchmark.

## Commercial fleets

## Stratification of the French fleet

The existence of official French statistics by vessel and trip (at least for the recent ten years), allows to stratify the whole fleet in order to propose homogeneous pools for commercial tuning fleets. Spatio-temporal variability of fishing power should also be performed aiming to evaluate the effect of different decommissioning plans throughout the time-series.

## Development of Ipue or cpue indicators based on VMS

Ireland has linked VMS and log book information for all vessels (See Gerritsen, 2009, WD 1). Using this data it is possible to develop lpue indicators at an appropriate spatial scale for mud patches with FU20-22. As the time-series develops these will become useful assessment input. This analysis should be extended to French and UK fleets fishing in FU20-22.

### 7.7.10 Management considerations

The average landings during 2000s have been stable with a slight increasing trend, but in 2009 total landings from the area declined and French global indicators on the stock remain unknown. However, various additional information such as mean sizes in landings, discard rate, abundances provided by UWTV survey suggest that there is little evidence for significant changes in the status of this stock.

ICES has repeatedly advised that management should be at a smaller scale than the ICES Subarea VII. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort are at the same scale as the resource.

The Nephrops fisheries target different areas, and Nephrops catches and landings show very different size structures. These fisheries also have differences in non-Nephrops bycatch composition. Cod, whiting, and to a lesser extent haddock are the main bycatch species.

Discarding of small Nephrops is substantial. The discard rate seems to have notably fluctuated between fleets or years. This shows that trawls currently used to target Nephrops are not technically adapted to select marketable Nephrops. The calculation of
the discard rate may be impacted by the upwards trend of tailed individuals in landings. Discarding of other fish species is also a problem in Nephrops fishery.

The French trawlers showed an overall decline during the last decade. It should be substantial to examine the evolution of the French fishing effort (decommissioning schemes associated to constraints linked to fuel prices). Irish fleet has also been impacted by European decommissioning plans.

In 2008, the lpue has increased for both fleets, but in 2009 Irish lpue changed downwards whereas the current lack of reliability for French official statistics does not allow for the estimation of French lpue indices for 2009. All lpue values over the whole time-series have not been corrected to take into account changing fishing power of fishing practices.

In 2006 and 2007, mean sizes in landings for both fleets decreased and that could be induced by stronger recruitment abundance than previously. However, since 2008, the French sampling plan at auction has included tailed individuals and modified interpretation of the signal. Back-calculations on mean sizes including tails since their proportion became significant (end of 1990s) were performed, but are not currently benchmarked.

Effort of Irish vessels is directed mainly in the Smalls ground which has high densities of small Nephrops. Currently, French effort is directed towards other grounds such as the Labbadie where the substrate is more heterogeneous and the mean size of Nephrops is significantly larger. There have been some changes in these patterns the over time. In 2009, Irish effort on the Labadie has increased and effort in the Smalls has declined (Gerritsen, 2010, WD 4).

### 7.7.11 References

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Table 7.7.1. Nephrops FU 20-22 (Celtic Sea). Total and by country nominal landings (t) in Division VIIfgh as used by WG.

| Year | France | Rep. of Ireland | UK | Other Countries ${ }^{1}$ | Total reported | Unallocated Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 3667 |  |  |  |  |  |  |
| 1984 | 3653 |  |  |  |  |  |  |
| 1985 | 3599 |  |  |  |  |  |  |
| 1986 | 2638 |  |  |  |  |  |  |
| 1987 | 3080 | 329 |  |  |  |  |  |
| 1988 | 2926 | 239 |  |  |  |  |  |
| 1989 | 3221 | 784 |  |  |  |  |  |
| 1990 | 3762 | 528 |  |  |  |  |  |
| 1991 | 2651 | 644 |  |  |  |  |  |
| 1992 | 3415 | 750 |  |  |  |  |  |
| 1993 | 3815 | 770 | 63 | 0 | 4648 | -274 | 4374 |
| 1994 | 3658 | 1415 | 68 | 2 | 5143 | -274 | 4869 |
| 1995 | 3803 | 1575 | 125 | 2 | 5505 | -282 | 5223 |
| 1996 | 3363 | 1377 | 86 | 2 | 4828 | -217 | 4611 |
| 1997 | 2589 | 1552 | 95 | 4 | 4240 | -213 | 4027 |
| 1998 | 2241 | 1619 | 64 | 1 | 3925 | -90 | 3835 |
| 1999 | 2078 | 824 | 41 | 0 | 2943 | -78 | 2865 |
| 2000 | 2848 | 1793 | 47 | 1 | 4689 | -44 | 4545 |
| 2001 | 2626 | 2123 | 21 | 1 | 4771 | -33 | 4738 |
| 2002 | 3154 | 1496 | 15 | 8 | 4673 | -50 | 4623 |
| 2003 | 3595 | 1388 | 19 | N/A | 5002 | 0 | 5002 |
| 2004 | 2605 | 1627 | 36 | N/A | 4268 | 0 | 4268 |
| 2005 | 2502 | 2391 | 53 | N/A | 4946 | 0 | 4946 |
| 2006 | 2368 | 1864 | 32 | N/A | 4264 | 0 | 4264 |
| 2007 | 2033 | 3213 | 47 | 6 | 5299 | 0 | 5299 |
| 2008 | 2348 | 3411 | 242 | N/A | 6001 | 0 | 6001 |
| 2009 | 2156 | 2844 | 359 | N/A | 5359 | 0 | 5359 |

${ }^{1}$ Other countries include Belgium.

Table 7.7.2. Nephrops FU 20-22 (Celtic Sea). Nominal landings (t) by quarter in Division VIIfgh as used by WG.

| year | French trawlers |  |  |  |  | Irish trawlers |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Q1 | Q2 | Q3 | Q4 | Total | Q1 | Q2 | Q3 | Q4 | Total |
| 1987 | 759 | 941 | 972 | 409 | 3080 |  |  |  |  | 329 |
| 1988 | 547 | 1065 | 683 | 631 | 2926 |  |  |  |  | 239 |
| 1989 | 411 | 1493 | 838 | 480 | 3221 |  |  |  |  | 784 |
| 1990 | 482 | 1765 | 1229 | 287 | 3762 |  |  |  |  | 528 |
| 1991 | 500 | 1245 | 518 | 388 | 2652 |  |  |  |  | 644 |
| 1992 | 681 | 992 | 1064 | 678 | 3415 |  |  |  |  | 750 |
| 1993 | 972 | 1598 | 742 | 504 | 3815 |  |  |  |  | 770 |
| 1994 | 541 | 1303 | 1052 | 762 | 3658 |  |  |  |  | 1415 |
| 1995 | 693 | 1631 | 876 | 604 | 3803 | 193 | 1137 | 109 | 136 | 1575 |
| 1996 | 674 | 1437 | 728 | 523 | 3363 | 268 | 714 | 330 | 66 | 1377 |
| 1997 | 460 | 1028 | 683 | 417 | 2589 | 249 | 971 | 196 | 136 | 1552 |
| 1998 | 642 | 881 | 456 | 262 | 2241 | 351 | 952 | 264 | 52 | 1619 |
| 1999 | 479 | 447 | 606 | 546 | 2078 | 214 | 184 | 105 | 321 | 824 |
| 2000 | 598 | 1261 | 743 | 246 | 2848 | 420 | 1154 | 149 | 71 | 1793 |
| 2001 | 422 | 879 | 667 | 658 | 2626 | 456 | 843 | 317 | 508 | 2123 |
| 2002 | 479 | 1211 | 823 | 641 | 3154 | 167 | 557 | 408 | 363 | 1496 |
| 2003 | 533 | 1401 | 1187 | 474 | 3595 | 202 | 519 | 478 | 190 | 1388 |
| 2004 | 496 | 981 | 677 | 452 | 2605 | 234 | 685 | 341 | 367 | 1627 |
| 2005 | 628 | 909 | 537 | 428 | 2502 | 491 | 1390 | 233 | 277 | 2391 |
| 2006 | 486 | 1024 | 563 | 295 | 2368 | 354 | 978 | 233 | 299 | 1864 |
| 2007 | 294 | 966 | 423 | 350 | 2033 | 416 | 1331 | 415 | 1051 | 3213 |
| 2008 | 450 | 794 | 681 | 424 | 2348 | 493 | 1589 | 600 | 728 | 3411 |
| 2009 | 534 | 890 | 489 | 244 | 2156 | 932 | 1186 | 529 | 197 | 2844 |

Table 7.7.3. Nephrops in VIIfgh. Length distribution of landings by country in 2002. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3.

- The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).
- The Irish data reported from the whole MA M (See Stock Annex).

| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails |  |
| 17 |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |  |
| 18 |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |  |
| 19 |  |  | 4 |  |  | 5 |  |  |  |  | 1 | 24 |  | 2 | 33 |
| 20 |  |  | 13 |  |  | 6 |  |  |  |  | 2 | 126 |  | 3 | 145 |
| 21 |  | 1 | 37 |  |  | 4 |  |  |  |  | 4 | 172 |  | 5 | 213 |
| 22 |  | 1 | 72 |  |  | 17 |  |  |  |  | 7 | 564 |  | 8 | 653 |
| 23 |  | 2 | 124 |  |  | 85 |  |  | 6 |  | 11 | 1124 |  | 14 | 1340 |
| 24 |  | 4 | 236 |  | 1 | 136 |  |  | 67 | 81 | 77 | 1804 | 81 | 83 | 2243 |
| 25 |  | 8 | 421 |  | 2 | 216 |  |  | 75 |  | 29 | 1533 |  | 39 | 2245 |
| 26 |  | 15 | 538 |  | 3 | 245 |  | 1 | 182 |  | 47 | 1495 |  | 66 | 2459 |
| 27 |  | 29 | 778 |  | 6 | 326 |  | 2 | 202 |  | 77 | 1110 |  | 113 | 2417 |
| 28 |  | 54 | 760 | 83 | 69 | 577 |  | 4 | 607 |  | 126 | 1516 | 83 | 254 | 3459 |
| 29 | 21 | 118 | 639 |  | 19 | 776 |  | 10 | 470 |  | 347 | 1220 | 21 | 494 | 3104 |
| 30 | 41 | 223 | 510 |  | 35 | 741 |  | 22 | 1125 | 242 | 685 | 1107 | 283 | 966 | 3483 |
| 31 | 47 | 395 | 589 |  | 65 | 1075 |  | 49 | 1685 | 242 | 733 | 1284 | 289 | 1242 | 4632 |
| 32 | 132 | 439 | 565 |  | 119 | 1199 |  | 108 | 1558 | 242 | 674 | 1002 | 375 | 1340 | 4325 |
| 33 | 140 | 459 | 453 | 83 | 278 | 1624 | 37 | 266 | 1551 | 404 | 725 | 995 | 664 | 1729 | 4624 |
| 34 | 236 | 523 | 419 | 122 | 879 | 1654 | 165 | 830 | 1455 | 404 | 739 | 753 | 927 | 2971 | 4281 |
| 35 | 366 | 609 | 326 | 540 | 1521 | 1654 | 401 | 1471 | 1152 | 678 | 866 | 782 | 1985 | 4466 | 3913 |
| 36 | 503 | 678 | 256 | 995 | 2072 | 1376 | 1125 | 1763 | 599 | 601 | 776 | 512 | 3223 | 5288 | 2742 |
| 37 | 648 | 744 | 221 | 1541 | 2279 | 1361 | 706 | 1360 | 711 | 823 | 905 | 412 | 3718 | 5288 | 2705 |
| 38 | 797 | 806 | 198 | 1603 | 2133 | 1156 | 1603 | 1752 | 580 | 1146 | 1083 | 526 | 5150 | 5774 | 2460 |
| 39 | 847 | 801 | 198 | 2230 | 2385 | 820 | 1463 | 1490 | 341 | 824 | 830 | 270 | 5364 | 5505 | 1628 |
| 40 | 1078 | 941 | 116 | 2901 | 2660 | 907 | 1466 | 1309 | 313 | 1618 | 1368 | 270 | 7063 | 6278 | 1606 |
| 41 | 817 | 712 | 47 | 2757 | 2350 | 380 | 1028 | 888 | 249 | 1377 | 1139 | 171 | 5978 | 5088 | 847 |
| 42 | 1114 | 915 | 140 | 2365 | 1905 | 322 | 1186 | 953 | 207 | 669 | 566 | 156 | 5334 | 4338 | 825 |
| 43 | 509 | 427 | 12 | 2070 | 1582 | 249 | 781 | 626 | 129 | 836 | 662 | 85 | 4196 | 3297 | 474 |
| 44 | 604 | 489 | 47 | 1003 | 784 | 234 | 1076 | 835 | 129 | 771 | 618 | 28 | 3454 | 2726 | 438 |
| 45 | 352 | 286 | 23 | 1157 | 877 | 132 | 605 | 475 | 74 | 612 | 523 | 71 | 2727 | 2161 | 300 |
| 46 | 144 | 121 |  | 467 | 368 | 132 | 893 | 691 | 37 | 306 | 278 | 14 | 1811 | 1459 | 183 |
| 47 | 179 | 149 |  | 345 | 301 | 15 | 470 | 371 | 97 | 247 | 236 | 14 | 1241 | 1057 | 126 |
| 48 | 78 | 67 | 23 | 472 | 389 | 102 | 422 | 331 | 55 | 175 | 160 | 14 | 1147 | 947 | 195 |
| 49 | 87 | 74 | 12 | 133 | 123 | 59 | 202 | 164 | 37 | 55 | 58 | 14 | 477 | 419 | 121 |
| 50 | 73 | 62 |  | 242 | 207 | 15 | 158 | 129 |  | 87 | 91 | 14 | 560 | 489 | 29 |


| CL | Q1 |  | Q2 |  | Q3 |  |  | Q4 |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 51 | 48 | 41 | 166 | 142 | 126 | 106 | 18 | 95 | 83 | 435 | 371 | 18 |
| 52 | 32 | 29 | 72 | 73 | 120 | 100 | 18 | 94 | 74 | 318 | 276 | 18 |
| 53 | 30 | 28 | 76 | 77 | 45 | 43 |  | 24 | 25 | 175 | 172 |  |
| 54 | 31 | 29 | 57 | 57 | 65 | 54 | 18 | 23 | 24 | 176 | 164 | 18 |
| 55 | 24 | 24 | 53 | 53 | 99 | 80 | 18 | 17 | 17 | 192 | 174 | 18 |
| 56 | 18 | 18 | 40 | 41 | 19 | 18 |  | 8 | 9 | 85 | 85 |  |
| 57 | 11 | 11 | 42 | 42 | 9 | 9 | 18 | 15 | 15 | 77 | 78 | 18 |
| 58 | 11 | 11 | 23 | 23 | 8 | 8 | 18 |  |  | 42 | 42 | 18 |
| 59 | 10 | 10 | 12 | 12 | 2 | 2 |  | 1 | 1 | 25 | 26 |  |
| 60 | 12 | 13 | 14 | 14 | 7 | 6 | 18 | 1 | 1 | 34 | 34 | 18 |
| 61 | 3 | 3 | 18 | 18 | 7 | 7 |  | 1 | 1 | 28 | 28 |  |
| 62 | 4 | 4 | 20 | 21 | 1 | 1 |  | 1 | 1 | 26 | 26 |  |
| 63 | 2 | 2 |  |  | 1 | 1 |  | 8 | 8 | 11 | 11 |  |
| 64 | 2 | 2 |  |  |  |  |  | 1 | 1 | 2 | 2 |  |
| 65 | 2 | 2 |  |  | 1 | 1 |  |  |  | 3 | 3 |  |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 | 1 | 1 |  |  | 1 | 1 |  |  |  | 2 | 2 |  |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |

Total 9056103817774217032398517600142931633413821127321470619184577836540658378

Table 7.7.4. Nephrops in VIIfgh. Length distribution of landings by country in 2003. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3.

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).


| CL | Q1 |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  | no <br> tails | tails |  | no <br> tails | tails |  | $\begin{aligned} & \text { no } \\ & \text { tails } \end{aligned}$ | tails |  | no tails | tails |  |
| 52 | 42 | 44 | 119 | 123 | 10 | 90 | 91 |  | 57 | 59 | 14 | 308 | 317 | 24 |
| 53 | 25 | 26 | 93 | 96 |  | 54 | 55 |  | 27 | 28 |  | 199 | 204 |  |
| 54 | 12 | 13 | 86 | 89 |  | 18 | 18 |  | 9 | 9 |  | 126 | 129 |  |
| 55 | 25 | 26 | 40 | 41 |  | 9 | 9 |  | 21 | 21 |  | 94 | 97 |  |
| 56 | 10 | 10 | 33 | 34 |  | 36 | 36 |  | 3 | 3 |  | 82 | 84 |  |
| 57 | 10 | 10 | 27 | 27 | 10 | 36 | 36 |  | 3 | 3 |  | 75 | 77 | 10 |
| 58 | 5 | 5 | 20 | 20 |  |  |  |  |  |  |  | 25 | 26 |  |
| 59 | 2 | 3 | 13 | 14 |  | 9 | 9 |  |  |  |  | 25 | 25 |  |
| 60 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 61 |  |  | 7 | 7 |  |  |  |  |  |  |  | 7 | 7 |  |
| 62 | 5 | 5 |  |  |  |  |  |  |  |  |  | 5 | 5 |  |
| 63 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 64 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 65 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Total 8424896312429229772550223767229782604322500858195149258629597002167953

Table 7.7.5. Nephrops in VIIfgh. Length distribution of landings by country in 2004. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3.

- The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).
- The missing Irish data of the 1st and 4th quarters were calculated by likelihood function as explained (Stock Annex).

| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |
| 18 |  |  | 3 |  |  |  |  |  |  |  |  | 2 |  |  | 6 |
| 19 |  |  | 16 |  |  |  |  |  |  |  |  | 4 |  |  | 20 |
| 20 |  |  | 30 |  |  | 1 |  |  | 1 |  |  | 8 |  | 1 | 40 |
| 21 |  |  | 46 |  |  | 11 |  | 1 | 1 |  |  | 19 |  | 1 | 77 |
| 22 |  |  | 69 |  |  | 8 |  | 1 |  |  |  | 57 |  | 3 | 134 |
| 23 |  | 1 | 108 |  | 1 | 25 |  | 3 | 4 |  | 1 | 107 |  | 5 | 245 |
| 24 |  | 1 | 160 |  | 1 | 100 |  | 5 | 13 |  | 1 | 207 |  | 9 | 480 |
| 25 |  | 3 | 213 |  | 1 | 189 |  | 11 | 37 |  | 2 | 368 |  | 17 | 806 |
| 26 |  | 5 | 298 |  | 3 | 445 |  | 21 | 107 |  | 4 | 565 |  | 32 | 1414 |
| 27 |  | 9 | 390 |  | 4 | 576 |  | 40 | 286 |  | 6 | 799 |  | 59 | 2052 |
| 28 |  | 16 | 443 |  | 8 | 703 |  | 78 | 699 |  | 10 | 1091 |  | 112 | 2935 |
| 29 |  | 29 | 537 |  | 13 | 1010 |  | 151 | 1126 |  | 17 | 1360 |  | 211 | 4034 |
| 30 |  | 53 | 680 |  | 23 | 1398 |  | 293 | 1652 |  | 30 | 1521 |  | 399 | 5251 |
| 31 |  | 97 | 737 |  | 39 | 1960 | 73 | 939 | 1798 |  | 50 | 1563 | 73 | 1125 | 6058 |
| 32 | 80 | 466 | 783 | 64 | 110 | 2487 | 254 | 1281 | 1606 |  | 84 | 1542 | 398 | 1942 | 6417 |
| 33 | 321 | 727 | 800 | 64 | 159 | 2862 | 363 | 1141 | 1403 |  | 143 | 1386 | 748 | 2170 | 6451 |
| 34 | 351 | 842 | 745 |  | 388 | 3030 | 327 | 992 | 1337 | 161 | 337 | 1144 | 838 | 2560 | 6256 |
| 35 | 728 | 993 | 633 | 191 | 631 | 2293 | 689 | 1188 | 988 | 183 | 633 | 908 | 1792 | 3445 | 4823 |
| 36 | 618 | 819 | 553 | 318 | 1231 | 1901 | 1161 | 1323 | 708 | 688 | 1120 | 738 | 2785 | 4492 | 3900 |
| 37 | 763 | 811 | 443 | 1080 | 1753 | 1698 | 871 | 961 | 449 | 1009 | 1245 | 544 | 3723 | 4770 | 3134 |
| 38 | 827 | 770 | 373 | 1080 | 1749 | 1299 | 1161 | 986 | 353 | 596 | 821 | 397 | 3664 | 4326 | 2422 |
| 39 | 537 | 499 | 298 | 1652 | 1728 | 797 | 798 | 664 | 225 | 688 | 695 | 297 | 3675 | 3586 | 1616 |
| 40 | 695 | 574 | 216 | 826 | 1006 | 498 | 980 | 740 | 134 | 573 | 550 | 223 | 3074 | 2869 | 1071 |
| 41 | 486 | 403 | 150 | 1525 | 1326 | 447 | 1161 | 836 | 135 | 573 | 498 | 162 | 3745 | 3063 | 893 |
| 42 | 612 | 481 | 105 | 1789 | 1401 | 249 | 762 | 544 | 82 | 688 | 532 | 118 | 3852 | 2958 | 554 |
| 43 | 516 | 405 | 68 | 837 | 683 | 161 | 726 | 508 | 57 | 575 | 428 | 79 | 2653 | 2023 | 365 |
| 44 | 461 | 366 | 41 | 1218 | 885 | 74 | 635 | 447 | 59 | 392 | 289 | 59 | 2706 | 1988 | 233 |
| 45 | 470 | 364 | 31 | 1092 | 823 | 50 | 527 | 370 | 30 | 482 | 339 | 46 | 2571 | 1896 | 156 |
| 46 | 129 | 118 | 21 | 827 | 598 |  | 142 | 111 | 22 | 432 | 294 | 29 | 1530 | 1121 | 72 |
| 47 | 309 | 248 | 16 | 457 | 367 | 50 | 408 | 309 | 24 | 90 | 73 | 17 | 1264 | 998 | 106 |
| 48 | 178 | 166 | 11 | 661 | 569 | 25 | 278 | 225 | 11 | 182 | 135 | 14 | 1299 | 1095 | 61 |
| 49 | 178 | 166 | 9 | 352 | 319 | 25 | 282 | 229 | 11 | 123 | 101 | 6 | 935 | 814 | 51 |


| CL | Q1 |  | Q2 |  | Q3 |  |  |  | Q4 |  |  | Year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails |  |  | $\begin{aligned} & \text { no } \\ & \text { tails } \end{aligned}$ | tails |  |
| 50 | 125 | 120 | 5 | 395 | 360 |  | 149 | 155 | 5 | 69 | 63 | 4 | 739 | 697 | 14 |
| 51 | 149 | 143 | 4 | 193 | 197 |  | 145 | 151 | 3 | 54 | 56 | 3 | 541 | 547 | 10 |
| 52 | 117 | 118 | 2 | 215 | 219 |  | 126 | 131 | 3 | 58 | 59 | 3 | 516 | 527 | 7 |
| 53 | 81 | 81 | 2 | 204 | 208 |  | 114 | 106 | 8 | 81 | 83 | 2 | 479 | 478 | 12 |
| 54 | 60 | 60 | 2 | 129 | 131 |  | 37 | 39 | 3 | 61 | 63 | 2 | 287 | 293 | 6 |
| 55 | 60 | 60 |  | 64 | 66 |  | 37 | 39 | 3 | 48 | 49 | 3 | 209 | 214 | 6 |
| 56 | 36 | 37 |  | 54 | 55 |  | 37 | 39 |  | 36 | 37 | 3 | 164 | 167 | 3 |
| 57 | 26 | 26 |  | 54 | 55 |  | 37 | 39 | 16 | 17 | 18 | 3 | 134 | 137 | 19 |
| 58 | 18 | 18 |  | 11 | 11 |  | 26 | 27 |  | 12 | 12 | 3 | 66 | 68 | 3 |
| 59 | 3 | 3 |  | 32 | 33 |  | 4 | 4 | 5 | 10 | 10 | 3 | 48 | 49 | 8 |
| 60 | 3 | 3 |  |  |  |  | 15 | 15 |  | 6 | 6 | 1 | 23 | 24 | 1 |
| 61 |  |  |  |  |  |  | 15 | 15 |  | 2 | 2 | 1 | 17 | 17 | 1 |
| 62 |  |  |  |  |  |  | 11 | 12 |  |  |  |  | 11 | 12 |  |
| 63 |  |  |  |  |  |  | 4 | 4 |  |  |  |  | 4 | 4 |  |
| 64 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 65 |  |  |  |  |  |  |  |  |  | 2 | 2 |  | 2 | 2 |  |
| 66 |  |  |  |  |  |  |  |  | 3 |  |  |  |  |  | 3 |
| 67 |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |
| 68 |  |  |  |  |  |  |  |  |  | 2 | 2 | 1 | 2 | 2 | 1 |
| 69 |  |  |  |  |  |  |  |  | 3 |  |  |  |  |  | 3 |
| 70 |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |
| 71 |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  | 1 |
| 72 |  |  |  |  |  |  |  |  | 3 |  |  |  |  |  | 3 |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total 89381009990421538117152243711235415173134117892890315412445655132762236 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.7.6. Nephrops in VIIfgh. Length distribution of landings by country in 2005. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3 .

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).

| CL | Q1 |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no tails |  |  | no <br> tails |  |  | no tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  | 3 |  |  |  |  |  |  |  |  | 3 |
| 19 |  |  | 10 |  |  | 7 |  |  |  |  |  |  |  |  | 17 |
| 20 |  |  | 22 |  |  | 31 |  |  | 41 |  | 1 | 107 |  | 1 | 200 |
| 21 |  |  | 22 |  |  | 74 |  |  | 187 |  | 1 | 379 |  | 1 | 662 |
| 22 |  |  | 60 |  |  | 79 |  |  | 226 |  | 1 | 422 |  | 1 | 787 |
| 23 |  | 1 | 150 |  |  | 209 |  |  | 343 | 1 | 2 | 811 | 1 | 3 | 1513 |
| 24 |  | 1 | 543 |  |  | 446 |  |  | 759 |  | 3 | 1506 |  | 4 | 3253 |
| 25 |  | 2 | 705 |  | 1 | 1035 |  |  | 1074 |  | 4 | 1874 |  | 7 | 4688 |
| 26 |  | 3 | 1303 |  | 1 | 1302 |  | 1 | 1143 | 8 | 12 | 3006 | 8 | 17 | 6754 |
| 27 | 9 | 12 | 1942 |  | 3 | 2227 |  | 1 | 1712 | 1 | 13 | 2404 | 10 | 29 | 8285 |
| 28 |  | 10 | 1878 |  | 5 | 2983 |  | 3 | 1897 | 2 | 21 | 1753 | 2 | 39 | 8510 |
| 29 |  | 18 | 2122 |  | 9 | 4281 |  | 6 | 1610 | 1 | 32 | 1296 | 1 | 64 | 9310 |
| 30 | 9 | 38 | 2281 |  | 16 | 5134 |  | 11 | 1404 | 4 | 59 | 1059 | 13 | 125 | 9878 |
| 31 |  | 57 | 2427 |  | 30 | 6639 |  | 23 | 1150 | 21 | 104 | 1048 | 21 | 214 | 11264 |
| 32 | 70 | 153 | 2056 |  | 56 | 7014 | 8 | 52 | 575 | 70 | 264 | 631 | 148 | 525 | 10276 |
| 33 | 44 | 426 | 1312 | 10 | 109 | 6247 | 18 | 105 | 709 | 162 | 506 | 491 | 233 | 1146 | 8759 |
| 34 | 131 | 573 | 1436 |  | 190 | 4688 | 58 | 628 | 439 | 471 | 867 | 624 | 660 | 2257 | 7187 |
| 35 | 289 | 792 | 1101 | 69 | 768 | 4429 | 196 | 833 | 169 | 769 | 1163 | 247 | 1323 | 3556 | 5945 |
| 36 | 464 | 876 | 688 | 223 | 1296 | 3546 | 297 | 948 | 140 | 1076 | 1322 | 322 | 2060 | 4443 | 4696 |
| 37 | 525 | 805 | 553 | 429 | 1455 | 1916 | 515 | 944 | 151 | 1188 | 1271 | 123 | 2656 | 4475 | 2743 |
| 38 | 578 | 752 | 557 | 483 | 1334 | 1985 | 558 | 852 | 62 | 1109 | 1064 | 192 | 2728 | 4003 | 2796 |
| 39 | 814 | 822 | 459 | 598 | 1134 | 1343 | 761 | 822 | 31 | 934 | 817 | 178 | 3106 | 3594 | 2011 |
| 40 | 658 | 637 | 379 | 615 | 924 | 659 | 696 | 652 | 31 | 731 | 599 | 69 | 2700 | 2813 | 1137 |
| 41 | 735 | 636 | 180 | 617 | 770 | 493 | 545 | 468 | 16 | 589 | 451 | 41 | 2487 | 2325 | 730 |
| 42 | 780 | 632 | 99 | 744 | 707 | 370 | 493 | 388 | 75 | 415 | 316 | 27 | 2432 | 2043 | 573 |
| 43 | 570 | 454 | 159 | 588 | 529 | 110 | 412 | 310 | 23 | 450 | 319 | 14 | 2021 | 1613 | 305 |
| 44 | 613 | 473 | 99 | 598 | 479 | 27 | 276 | 212 | 60 | 288 | 212 |  | 1775 | 1376 | 186 |
| 45 | 547 | 418 |  | 746 | 544 | 27 | 247 | 192 |  | 271 | 198 | 14 | 1812 | 1352 | 41 |
| 46 | 520 | 402 | 80 | 701 | 493 | 82 | 161 | 134 |  | 182 | 139 |  | 1563 | 1169 | 163 |
| 47 | 400 | 312 |  | 752 | 513 | 27 | 199 | 164 |  | 135 | 110 |  | 1486 | 1099 | 27 |
| 48 | 258 | 218 |  | 757 | 511 |  | 158 | 135 | 68 | 75 | 66 |  | 1248 | 931 | 68 |
| 49 | 271 | 238 |  | 677 | 461 |  | 177 | 134 |  | 49 | 48 |  | 1174 | 881 |  |
| 50 | 241 | 220 |  | 698 | 489 | 41 | 302 | 226 |  | 34 | 35 |  | 1275 | 969 | 41 |
| 51 | 263 | 239 |  | 476 | 350 |  | 271 | 203 |  | 40 | 42 |  | 1051 | 833 |  |


| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F | IRL |
|  | no tails |  |  | no <br> tails | tails |  | no <br> tails |  |  | no <br> tails |  |  | no tails | tails |
| 52 | 179 | 171 |  | 349 | 277 |  | 215 | 165 |  | 21 | 22 |  | 764 | 635 |
| 53 | 153 | 139 |  | 332 | 262 |  | 198 | 144 |  | 23 | 24 |  | 707 | 569 |
| 54 | 101 | 101 |  | 241 | 193 |  | 181 | 133 |  | 20 | 20 |  | 543 | 448 |
| 55 | 89 | 88 |  | 193 | 167 |  | 205 | 149 |  | 16 | 16 |  | 502 | 420 |
| 56 | 50 | 51 |  | 132 | 114 |  | 85 | 64 |  | 9 | 9 |  | 276 | 238 |
| 57 | 58 | 56 |  | 140 | 106 |  | 73 | 56 |  | 9 | 9 |  | 280 | 228 |
| 58 | 33 | 33 |  | 64 | 53 |  | 68 | 50 |  | 4 | 5 |  | 169 | 141 |
| 59 | 31 | 32 |  | 48 | 41 |  | 48 | 35 |  | 5 | 5 |  | 133 | 113 |
| 60 | 15 | 15 |  | 8 | 8 |  | 13 | 14 |  | 4 | 4 |  | 39 | 41 |
| 61 | 15 | 15 |  | 9 | 9 |  | 18 | 13 |  | 1 | 1 |  | 43 | 39 |
| 62 | 3 | 3 |  | 5 | 5 |  | 4 | 7 |  |  |  |  | 11 | 15 |
| 63 | 3 | 3 |  | 3 | 3 |  | 10 | 8 |  | 1 | 1 |  | 17 | 15 |
| 64 |  |  |  |  |  |  | 1 | 2 |  |  |  |  | 1 | 2 |
| 65 |  |  |  | 2 | 2 |  | 1 | 2 |  |  |  |  | 2 | 3 |
| 66 |  |  |  | 2 | 2 |  | 1 | 2 |  |  |  |  | 3 | 4 |
| 67 |  |  |  |  |  |  | 1 | 2 |  |  |  |  | 1 | 2 |
| 68 |  |  |  |  |  |  | 1 | 2 |  |  |  |  | 1 | 2 |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  | 1 | 2 |  |  |  |  | 1 | 2 |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  | 1 | 1 |  |  |  |  | 1 | 1 |
| 73 |  |  |  |  |  |  |  | 1 |  |  |  |  |  | 1 |
| 74 |  |  |  |  |  |  |  | 1 |  |  |  |  |  | 1 |
| 75 |  |  |  |  |  |  | 1 | 3 |  |  |  |  | 1 | 3 |

Total 951910928226201130714417574557474930414093919010181186393749144830112807

Table 7.7.7. Nephrops in VIIfgh. Length distribution of landings by country in 2006. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3 .

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).


| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  |  | no <br> tails | tails |  | $\begin{aligned} & \text { no } \\ & \text { tails } \end{aligned}$ | tails |  | $\begin{aligned} & \text { no } \\ & \text { tails } \end{aligned}$ |  |  | no <br> tails | tails |  |
| 52 | 68 | 68 |  | 156 | 127 | 13 | 70 | 63 |  | 19 | 18 |  | 313 | 276 | 13 |
| 53 | 62 | 64 |  | 114 | 101 |  | 46 | 52 |  | 10 | 11 |  | 231 | 227 |  |
| 54 | 42 | 44 |  | 72 | 69 |  | 42 | 39 |  | 9 | 10 |  | 166 | 161 |  |
| 55 | 34 | 35 |  | 63 | 59 |  | 27 | 28 |  | 10 | 10 |  | 134 | 133 |  |
| 56 | 33 | 35 |  | 39 | 41 |  | 23 | 24 |  | 8 | 9 |  | 105 | 108 |  |
| 57 | 29 | 30 |  | 38 | 39 |  | 13 | 14 |  | 5 | 5 |  | 85 | 87 |  |
| 58 | 17 | 18 |  | 38 | 39 |  | 12 | 12 |  | 5 | 5 |  | 71 | 74 |  |
| 59 | 11 | 11 | 13 | 26 | 27 |  | 8 | 9 |  | 3 | 4 |  | 49 | 50 | 13 |
| 60 | 7 | 7 |  | 15 | 15 |  | 12 | 12 |  | 2 | 2 |  | 36 | 37 |  |
| 61 | 4 | 4 |  | 10 | 11 |  | 6 | 6 |  | 1 | 1 |  | 21 | 22 |  |
| 62 | 3 | 3 |  | 3 | 3 |  | 4 | 4 |  | 1 | 1 |  | 10 | 11 |  |
| 63 | 1 | 1 |  |  |  |  | 1 | 1 |  | 1 | 1 |  | 3 | 3 |  |
| 64 | 2 | 2 |  | 2 | 2 |  | 2 | 2 |  |  |  |  | 7 | 7 |  |
| 65 |  |  |  | 1 | 1 |  | 1 | 1 |  |  |  |  | 2 | 2 |  |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 69 |  | 1 |  |  |  |  |  |  |  |  |  |  |  | 1 |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Total 820993362354317796205175054610060126751358962496959203284231549487108006

Table 7.7.8. Nephrops in VIIfgh. Length distribution of landings by country in 2007. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3.

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (simulation of hand-sorting s-shaped curve vs. CL: see Stock Annex).


| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails |  | no tails | tails |  | no <br> tails | tails |  |
| 52 | 34 | 31 |  | 70 | 62 |  | 19 | 20 |  | 20 | 18 |  | 142 | 132 |  |
| 53 | 22 | 21 |  | 39 | 41 |  | 11 | 12 |  | 25 | 19 | 24 | 98 | 93 | 24 |
| 54 | 18 | 17 |  | 21 | 22 |  | 9 | 9 |  | 27 | 19 |  | 76 | 67 |  |
| 55 | 19 | 18 |  | 17 | 18 |  | 8 | 8 |  | 6 | 6 |  | 50 | 50 |  |
| 56 | 9 | 9 |  | 18 | 19 |  | 5 | 5 |  | 19 | 12 |  | 51 | 46 |  |
| 57 | 7 | 7 |  | 7 | 7 |  | 2 | 2 |  | 8 | 6 |  | 24 | 22 |  |
| 58 | 11 | 10 | - | 6 | 6 | 14 | 2 | 2 |  | 2 | 2 |  | 21 | 20 | 14 |
| 59 | 4 | 4 |  | 5 | 5 |  |  |  |  | 1 | 1 |  | 10 | 10 |  |
| 60 | 5 | 5 |  | 6 | 6 |  | 1 | 1 |  | 2 | 2 |  | 13 | 13 |  |
| 61 | 2 | 2 |  | 5 | 5 |  | 1 | 1 |  | 1 | 1 |  | 8 | 9 |  |
| 62 | 2 | 2 |  | 3 | 4 |  | 1 | 1 |  |  |  |  | 7 | 7 |  |
| 63 | 1 | 1 |  | 2 | 2 |  |  |  |  |  |  |  | 3 | 4 |  |
| 64 |  |  |  | 1 | 1 |  |  |  |  |  |  |  | 2 | 2 |  |
| 65 |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 |  |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Total 52966293335201835421736842628897103222755362567366562533880345716201588

Table 7.7.9. Nephrops in VIIfgh. Length distribution of landings by country in 2008. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3.

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (as performed since WGCSE 2009).

| CL | Q1 |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails |  |  | no <br> tails | tails |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 20 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 21 |  |  |  |  |  | 28 |  |  |  |  |  |  |  |  | 28 |
| 22 |  |  |  |  |  | 296 |  |  |  |  |  |  |  |  | 296 |
| 23 |  |  |  |  |  | 651 |  |  | 69 |  |  | 539 |  |  | 1258 |
| 24 |  |  |  |  |  | 1475 |  |  | 410 |  |  | 1736 |  |  | 3621 |
| 25 |  |  | 18 |  |  | 2557 |  |  | 913 |  |  | 3494 |  |  | 6981 |
| 26 |  |  | 958 |  | 27 | 4475 |  | 22 | 1136 |  |  | 5829 |  | 49 | 12397 |
| 27 |  |  | 1011 |  | 82 | 5408 |  | 22 | 1782 |  |  | 1578 |  | 104 | 9779 |
| 28 |  | 26 | 3759 |  | 218 | 6541 |  | 89 | 1582 |  | 10 | 2856 |  | 343 | 14738 |
| 29 | 6 | 4 | 3033 |  | 463 | 6436 | 10 | 72 | 2256 | 6 | 43 | 1777 | 22 | 582 | 13502 |
| 30 | 6 | 162 | 3336 | 12 | 742 | 7257 |  | 245 | 2116 |  | 108 | 1878 | 18 | 1256 | 14588 |
| 31 | 19 | 275 | 980 | 13 | 1042 | 7312 |  | 467 | 2969 | 18 | 167 | 1419 | 50 | 1951 | 12680 |
| 32 | 38 | 497 | 1087 | 61 | 1774 | 6648 | 20 | 989 | 3241 | 55 | 307 | 1460 | 174 | 3567 | 12436 |
| 33 | 89 | 752 | 1319 | 280 | 1527 | 4916 | 30 | 1372 | 3063 | 146 | 488 | 1520 | 544 | 4140 | 10817 |
| 34 | 247 | 1058 | 1123 | 536 | 1789 | 4829 | 181 | 1629 | 2363 | 273 | 721 | 1698 | 1236 | 5198 | 10013 |
| 35 | 438 | 977 | 1462 | 925 | 1818 | 4573 | 441 | 1720 | 1221 | 450 | 817 | 1939 | 2253 | 5332 | 9194 |
| 36 | 554 | 1167 | 1123 | 1448 | 1993 | 3000 | 941 | 2116 | 1383 | 753 | 979 | 1219 | 3697 | 6254 | 6725 |
| 37 | 668 | 920 | 677 | 1692 | 1596 | 2042 | 1422 | 1589 | 718 | 863 | 897 | 900 | 4645 | 5001 | 4337 |
| 38 | 647 | 751 | 659 | 1814 | 1383 | 1224 | 1682 | 1525 | 666 | 1087 | 1032 | 999 | 5231 | 4690 | 3548 |
| 39 | 669 | 567 | 356 | 1583 | 1242 | 915 | 2063 | 1434 | 244 | 844 | 828 | 780 | 5159 | 4071 | 2294 |
| 40 | 597 | 444 | 339 | 1558 | 1148 | 562 | 1462 | 965 | 213 | 911 | 750 | 600 | 4528 | 3306 | 1713 |
| 41 | 654 | 465 | 267 | 1418 | 946 | 378 | 1382 | 856 | 282 | 772 | 619 | 679 | 4226 | 2886 | 1606 |
| 42 | 560 | 383 | 178 | 1027 | 671 | 393 | 1052 | 595 | 182 | 744 | 566 | 439 | 3383 | 2215 | 1192 |
| 43 | 576 | 367 | 89 | 1044 | 607 | 267 | 703 | 368 | 91 | 521 | 378 | 280 | 2845 | 1720 | 726 |
| 44 | 511 | 316 | 89 | 812 | 471 | 321 | 782 | 414 |  | 374 | 291 | 60 | 2480 | 1493 | 470 |
| 45 | 598 | 371 | 53 | 568 | 342 | 84 | 455 | 245 |  | 255 | 233 | 160 | 1876 | 1190 | 297 |
| 46 | 345 | 225 |  | 405 | 259 | 84 | 277 | 180 |  | 198 | 171 | 40 | 1225 | 835 | 123 |
| 47 | 290 | 206 |  | 219 | 151 |  | 184 | 112 |  | 118 | 123 | 40 | 812 | 593 | 40 |
| 48 | 209 | 144 |  | 201 | 173 | 41 | 105 | 76 |  | 84 | 62 | 40 | 600 | 456 | 81 |
| 49 | 102 | 74 |  | 128 | 97 | 167 | 100 | 76 |  | 65 | 50 | 40 | 395 | 298 | 207 |
| 50 | 117 | 84 |  | 93 | 81 | 125 | 55 | 45 |  | 44 | 36 | 40 | 308 | 247 | 165 |
| 51 | 49 | 39 |  | 56 | 56 | 41 | 74 | 60 |  | 50 | 37 | 20 | 229 | 192 | 61 |


| CL | Q1 |  |  | Q2 |  |  | Q3 |  |  | Q4 |  | Year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F | IRL | F |  | IRL |
|  | no tails |  |  | no <br> tails | tails |  | no <br> tails | tails |  | no <br> tails | tails | no <br> tails | tails |  |
| 52 | 28 | 25 |  | 47 | 40 | 41 | 30 | 30 |  | 17 | 14 | 120 | 109 | 41 |
| 53 | 36 | 29 |  | 28 | 28 |  | 23 | 23 |  | 14 | 12 | 102 | 92 |  |
| 54 | 11 | 11 |  | 21 | 21 |  | 16 | 16 |  | 6 | 16 | 55 | 65 |  |
| 55 | 13 | 11 |  | 17 | 17 |  | 12 | 12 |  | 3 | 3 | 46 | 43 |  |
| 56 | 8 | 8 |  | 12 | 12 |  | 7 | 7 |  | 1 | 1 | 28 | 28 |  |
| 57 | 12 | 10 |  | 7 | 7 |  | 5 | 5 |  | 2 | 2 | 27 | 24 |  |
| 58 | 14 | 12 |  | 4 | 4 |  | 1 | 1 |  | 1 | 1 | 20 | 17 |  |
| 59 | 4 | 4 |  | 3 | 3 |  | 1 | 1 |  |  |  | 8 | 8 |  |
| 60 | 1 | 1 |  | 3 | 3 |  | 1 | 1 |  |  |  | 4 | 4 |  |
| 61 |  |  |  | 1 | 1 |  |  |  |  |  |  | 2 | 2 |  |
| 62 |  |  |  | 1 | 1 |  |  |  |  |  |  | 1 | 1 |  |
| 63 |  |  |  | 1 | 1 |  |  |  |  |  |  | 1 | 1 |  |
| 64 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 65 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Total 8117103872191416039208367308613516173802690086769763340564634858365155956

Table 7.7.10. Nephrops in VIIfgh. Length distribution of landings by country in 2009. Quarterly and total values $\left(10^{3}\right)$. The reported size is the carapace length (CL). Conversion of CL to TS (total size) is done by multiplication by 3.3 .

The French data are presented by two ways: (1) Previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction). (2) Tails are included (as performed since WGCSE 2009).

| CL | Q1 |  | Q2 |  |  | Q3 |  |  | Q4 |  |  | Year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no <br> tails |  |  | no tails |  |  | no tails |  |  | no tails |  |  | no tails | tails |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 20 |  |  | 116 |  |  |  |  |  | 11 |  |  |  |  |  | 127 |
| 21 |  |  | 167 |  |  |  |  |  |  |  |  |  |  |  | 167 |
| 22 |  |  | 399 |  |  | 35 |  |  | 31 |  |  | 102 |  |  | 566 |
| 23 |  |  | 1017 |  |  | 217 |  |  | 103 |  |  | 306 |  |  | 1643 |
| 24 |  |  | 2582 |  |  | 505 |  |  | 364 |  |  | 756 |  |  | 4207 |
| 25 |  |  | 3963 |  |  | 1284 |  |  | 879 |  |  | 1279 |  |  | 7405 |
| 26 |  |  | 6524 |  |  | 1969 |  |  | 1536 |  |  | 1495 |  |  | 11525 |
| 27 |  |  | 5825 |  |  | 3351 |  |  | 2396 |  | 4 | 759 |  | 4 | 12331 |
| 28 |  |  | 4684 |  |  | 3619 |  | 14 | 2953 |  | 22 | 489 |  | 35 | 11744 |
| 29 |  |  | 5095 |  | 106 | 3889 |  | 14 | 2804 |  | 30 | 831 |  | 150 | 12619 |
| 30 |  | 15 | 3619 |  | 252 | 3852 |  | 151 | 2735 |  | 69 | 658 |  | 487 | 10865 |
| 31 |  | 166 | 2509 |  | 583 | 3759 |  | 329 | 1813 | 5 | 163 | 549 | 5 | 1241 | 8630 |
| 32 | 11 | 234 | 2044 |  | 769 | 3074 | 10 | 637 | 2361 | 9 | 152 | 754 | 31 | 1792 | 8234 |
| 33 | 34 | 309 | 1671 | 32 | 894 | 2872 | 41 | 736 | 1716 | 23 | 295 | 472 | 131 | 2233 | 6731 |
| 34 | 125 | 595 | 1799 | 205 | 1365 | 2222 | 10 | 705 | 1273 | 92 | 370 | 400 | 432 | 3035 | 5694 |
| 35 | 194 | 685 | 1285 | 488 | 1449 | 2003 | 249 | 985 | 1117 | 129 | 482 | 242 | 1059 | 3601 | 4647 |
| 36 | 479 | 991 | 1003 | 678 | 1759 | 1839 | 425 | 1011 | 774 | 267 | 434 | 417 | 1849 | 4196 | 4032 |
| 37 | 673 | 997 | 1119 | 1165 | 1828 | 1433 | 632 | 1027 | 603 | 345 | 454 | 242 | 2814 | 4306 | 3397 |
| 38 | 844 | 1048 | 1054 | 1714 | 1827 | 1369 | 902 | 967 | 502 | 419 | 442 | 181 | 3878 | 4283 | 3106 |
| 39 | 1072 | 1076 | 694 | 1885 | 1741 | 1339 | 912 | 780 | 380 | 524 | 444 | 157 | 4394 | 4040 | 2569 |
| 40 | 1028 | 911 | 411 | 1839 | 1542 | 808 | 1129 | 898 | 209 | 465 | 395 | 199 | 4461 | 3746 | 1627 |
| 41 | 935 | 790 | 823 | 1972 | 1383 | 724 | 987 | 644 | 236 | 410 | 329 | 48 | 4304 | 3145 | 1831 |
| 42 | 913 | 685 | 308 | 1575 | 1085 | 420 | 832 | 478 | 113 | 489 | 337 | 24 | 3808 | 2585 | 864 |
| 43 | 732 | 523 | 334 | 1438 | 968 | 288 | 837 | 524 | 175 | 345 | 244 |  | 3352 | 2259 | 797 |
| 44 | 703 | 555 | 154 | 1206 | 756 | 231 | 651 | 424 | 84 | 314 | 216 | 48 | 2875 | 1951 | 517 |
| 45 | 495 | 336 | 102 | 690 | 451 | 89 | 302 | 201 | 25 | 174 | 140 | 24 | 1660 | 1128 | 240 |
| 46 | 486 | 373 | 77 | 411 | 305 | 160 | 332 | 221 | 44 | 193 | 135 | 12 | 1422 | 1035 | 293 |
| 47 | 275 | 203 | 77 | 447 | 335 | 29 | 193 | 163 | 8 | 118 | 95 | 24 | 1033 | 796 | 137 |
| 48 | 233 | 196 | 102 | 147 | 127 | 43 | 136 | 107 |  | 63 | 52 | 24 | 579 | 482 | 169 |
| 49 | 142 | 118 |  | 175 | 156 | 29 | 139 | 110 |  | 67 | 52 | 12 | 523 | 435 | 40 |
| 50 | 77 | 73 |  | 101 | 88 | 43 | 113 | 79 | 8 | 31 | 29 |  | 321 | 268 | 51 |
| 51 | 37 | 52 |  | 97 | 90 | 29 | 38 | 34 |  | 20 | 20 |  | 192 | 196 | 29 |


| CL | Q1 |  |  | Q2 |  | Q3 |  |  | Q4 |  |  | Year |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (mm) | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL | F |  | IRL |
|  | no tails |  |  | no <br> tails | tails |  | no <br> tails |  |  | $\begin{aligned} & \text { no } \\ & \text { tails } \end{aligned}$ |  |  | no <br> tails | tails |  |
| 52 | 32 | 32 |  | 51 | 51 | 57 | 23 | 23 | 11 | 11 | 11 |  | 116 | 116 | 68 |
| 53 | 18 | 18 |  | 37 | 37 | 43 | 16 | 16 |  | 9 | 9 |  | 81 | 81 | 43 |
| 54 | 10 | 10 |  | 24 | 24 | 171 | 13 | 13 |  | 5 | 9 |  | 51 | 55 | 171 |
| 55 | 10 | 10 |  | 35 | 28 | 86 | 6 | 6 |  | 2 | 2 |  | 52 | 45 | 86 |
| 56 | 6 | 6 |  | 10 | 10 | 171 | 3 | 3 |  | 1 | 1 |  | 20 | 20 | 171 |
| 57 | 1 | 1 |  | 8 | 8 | 57 | 1 | 1 |  | 1 | 1 |  | 11 | 11 | 57 |
| 58 | 1 | 1 |  | 1 | 1 | 86 | 1 | 1 |  | 1 | 1 |  | 4 | 4 | 86 |
| 59 | 1 | 1 |  | 1 | 1 | 57 |  |  |  | 1 | 1 |  | 3 | 3 | 57 |
| 60 | 3 | 3 |  | 1 | 1 | 86 |  |  |  |  |  |  | 4 | 4 | 86 |
| 61 |  |  |  | 1 | 1 | 71 |  |  |  | 1 | 1 |  | 2 | 2 | 71 |
| 62 |  |  |  |  |  | 43 |  |  |  |  |  |  |  |  | 43 |
| 63 |  |  |  |  |  | 29 |  |  |  |  |  |  |  |  | 29 |
| 64 |  |  |  |  |  | 57 |  |  |  |  |  |  |  |  | 57 |
| 65 |  |  |  |  |  | 14 |  |  |  |  |  |  |  |  | 14 |
| 66 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 67 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  |  |  |  | 14 |  |  |  |  |  |  |  |  | 14 |
| 69 |  |  |  |  |  | 14 |  |  |  |  |  |  |  |  | 14 |
| 70 |  |  |  |  |  | 14 |  |  |  |  |  |  |  |  | 14 |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Total 956911013495571643320020425908933112992526345315441105053946747773127915

Table 7.7.11. Nephrops in FUs 20-22 Celtic Sea (VIIfgh) landings length distributions in 19871998. French trawlers.

| Landings CL mm/ | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 |  |  |  |  |  |  |  |  |  |  |  |  |
| 11 |  |  |  |  |  |  |  |  |  |  |  |  |
| 12 |  |  |  |  |  |  |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  |  |  |  |  |  |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |  |  |  |
| 20 |  |  |  |  |  |  |  |  |  |  |  |  |
| 21 |  | 57 |  | 7 |  |  |  |  |  |  |  |  |
| 22 |  |  |  |  |  |  | 38 |  |  |  |  |  |
| 23 |  | 53 |  | 36 |  |  |  |  | 43 |  |  |  |
| 24 |  | 106 |  | 57 |  | 30 |  |  | 43 |  |  |  |
| 25 | 24 | 289 |  |  | 14 |  | 85 |  | 86 |  |  |  |
| 26 | 88 | 309 |  | 29 | 53 | 60 | 19 | 12 | 109 | 15 |  |  |
| 27 | 149 | 490 |  | 143 | 34 | 111 | 84 | 23 | 644 | 20 | 15 |  |
| 28 | 684 | 1177 | 110 | 465 | 448 | 669 | 111 | 78 | 601 | 60 | 28 | 59 |
| 29 | 1104 | 3180 | 710 | 728 | 922 | 966 | 213 | 309 | 610 | 62 | 45 | 93 |
| 30 | 2030 | 4373 | 958 | 1241 | 1719 | 2139 | 393 | 631 | 1113 | 246 | 236 | 294 |
| 31 | 2317 | 7579 | 1804 | 2146 | 3047 | 3212 | 935 | 1113 | 1074 | 696 | 542 | 475 |
| 32 | 3640 | 8076 | 3103 | 2521 | 4057 | 4393 | 2253 | 2650 | 2486 | 1803 | 1220 | 1043 |
| 33 | 4449 | 8059 | 4294 | 4456 | 6036 | 6608 | 2468 | 3177 | 3203 | 2699 | 2144 | 1396 |
| 34 | 4312 | 8452 | 5210 | 5034 | 5804 | 6509 | 3757 | 4532 | 3129 | 4239 | 2186 | 2308 |
| 35 | 6179 | 6948 | 6479 | 6677 | 5721 | 7896 | 5213 | 6666 | 4870 | 6136 | 3608 | 3354 |
| 36 | 5691 | 5137 | 5914 | 5800 | 4591 | 8225 | 5941 | 5440 | 4339 | 5583 | 3827 | 3587 |
| 37 | 5479 | 5084 | 5281 | 5077 | 3959 | 8066 | 6026 | 6653 | 7127 | 6995 | 4262 | 4465 |
| 38 | 4940 | 3623 | 5931 | 6143 | 3797 | 7579 | 6784 | 6950 | 7141 | 7410 | 4804 | 4525 |
| 39 | 3870 | 2383 | 4832 | 5402 | 3091 | 5528 | 5667 | 4853 | 5497 | 5691 | 3619 | 3127 |
| 40 | 4622 | 2590 | 4843 | 4796 | 2772 | 3386 | 7263 | 5497 | 6493 | 5277 | 4918 | 4453 |
| 41 | 2482 | 2302 | 3636 | 3702 | 2216 | 2745 | 5349 | 4396 | 4044 | 4225 | 3062 | 2875 |
| 42 | 2695 | 2462 | 3675 | 4147 | 2218 | 2919 | 5485 | 4473 | 4433 | 4096 | 3414 | 2996 |
| 43 | 1994 | 1645 | 2371 | 3271 | 2110 | 2429 | 3652 | 3222 | 3257 | 3205 | 2725 | 2267 |
| 44 | 1275 | 1274 | 2165 | 3235 | 1793 | 1680 | 2415 | 2580 | 3403 | 2115 | 1849 | 2109 |
| 45 | 1590 | 1231 | 1999 | 2366 | 1550 | 1636 | 2732 | 2183 | 2142 | 2086 | 2288 | 1474 |
| 46 | 1265 | 988 | 1415 | 2066 | 1229 | 1222 | 1653 | 1348 | 1747 | 1183 | 1428 | 1014 |
| 47 | 1184 | 806 | 1151 | 1446 | 865 | 939 | 1604 | 1323 | 1635 | 1247 | 1021 | 1012 |
| 48 | 1182 | 778 | 858 | 1787 | 1057 | 966 | 1134 | 1204 | 1338 | 877 | 970 | 789 |
| 49 | 767 | 525 | 708 | 1277 | 766 | 738 | 950 | 898 | 816 | 747 | 603 | 433 |
| 50 | 834 | 437 | 565 | 809 | 527 | 576 | 981 | 969 | 972 | 702 | 733 | 420 |
| 51 | 571 | 307 | 511 | 692 | 437 | 406 | 489 | 639 | 743 | 504 | 353 | 274 |
| 52 | 668 | 353 | 447 | 786 | 403 | 278 | 612 | 571 | 770 | 510 | 372 | 253 |
| 53 | 526 | 260 | 315 | 477 | 303 | 303 | 365 | 395 | 635 | 389 | 286 | 157 |
| 54 | 268 | 205 | 253 | 387 | 236 | 191 | 344 | 462 | 448 | 294 | 198 | 110 |
| 55 | 391 | 111 | 148 | 204 | 128 | 171 | 276 | 364 | 262 | 197 | 110 | 109 |
| 56 | 150 | 107 | 156 | 95 | 121 | 96 | 162 | 191 | 152 | 141 | 54 | 76 |
| 57 | 129 | 85 | 118 | 90 | 48 | 74 | 93 | 110 | 176 | 116 | 81 | 41 |
| 58 | 55 | 49 | 96 | 91 | 73 | 68 | 83 | 154 | 124 | 56 | 36 | 28 |
| 59 | 92 | 33 | 74 | 31 | 12 | 48 | 93 | 68 | 49 | 22 | 8 | 7 |
| 60 | 52 | 4 | 26 | 26 | 17 | 24 | 47 | 71 | 69 | 17 | 23 | 13 |
| 61 | 7 | 4 | 22 | 8 |  | 11 | 19 | 22 | 22 | 5 | 8 |  |
| 62 | 11 | 10 | 7 | 21 | 7 | 9 | 25 | 9 | 29 | 20 | 3 |  |
| 63 | 6 |  | 12 |  | 1 |  | 5 | 12 | 13 | 2 |  | 2 |
| 64 |  |  | 5 |  |  |  |  |  |  |  |  |  |
| 65 | 16 | 4 | 5 |  |  |  | 6 | 2 | 3 |  |  |  |
| 66 |  |  |  |  |  | 2 |  | 2 |  |  |  |  |
| 67 | 6 |  |  |  |  |  |  |  |  |  |  |  |
| 68 |  |  | 5 |  |  |  |  |  |  |  |  |  |
| 69 |  |  |  |  |  |  |  |  |  |  |  |  |
| 70 |  |  |  |  |  |  |  |  |  |  |  |  |
| 71 |  |  |  |  |  |  |  |  |  |  |  |  |
| 72 |  |  |  |  |  |  |  | 2 |  |  |  |  |
| 73 |  |  |  |  |  |  |  |  |  |  |  |  |
| 74 |  |  |  |  |  |  |  |  |  |  |  |  |
| 75 |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 67794 | 81948 | 70215 | 77770 | 62182 | 82908 | 75824 | 74255 | 75892 | 69686 | 51080 | 45637 |
| Weights | 3080 | 2926 | 3221 | 3762 | 2652 | 3415 | 3815 | 3658 | 3803 | 3363 | 2589 | 2241 |

Table 7.7.12. Nephrops in FUs 20-22 Celtic Sea (VIIfgh) landings length distributions in 1999-2009. French trawlers.

Years 2008 and 2009: DLFs including tails are provided by sampling at auction. For previous years sampling involves only in entire Nephrops; DLF including tails are estimated by simulation (see Stock Annex).


Table 7.7.13a. Nephrops in FUs 20-22 Celtic Sea (VIIfgh). Landings-length distributions in 20022009. Irish trawlers.

| Landings |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CL mm/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 17 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 18 | 0 | 2 | 6 | 3 | 4 | 0 | 0 | 0 |
| 19 | 33 | 10 | 20 | 17 | 15 | 29 | 0 | 0 |
| 20 | 145 | 269 | 40 | 200 | 233 | 467 | 0 | 127 |
| 21 | 213 | 1068 | 77 | 662 | 504 | 601 | 28 | 167 |
| 22 | 653 | 1330 | 134 | 787 | 1123 | 2360 | 296 | 566 |
| 23 | 1340 | 2570 | 245 | 1513 | 2334 | 8020 | 1258 | 1643 |
| 24 | 2243 | 2139 | 480 | 3253 | 4554 | 13134 | 3621 | 4207 |
| 25 | 2245 | 2473 | 806 | 4688 | 6442 | 18142 | 6981 | 7405 |
| 26 | 2459 | 5815 | 1414 | 6754 | 10118 | 22705 | 12397 | 11525 |
| 27 | 2417 | 4370 | 2052 | 8285 | 12544 | 22221 | 9779 | 12331 |
| 28 | 3459 | 6644 | 2935 | 8510 | 13457 | 24482 | 14738 | 11744 |
| 29 | 3104 | 5574 | 4034 | 9310 | 12521 | 24017 | 13502 | 12619 |
| 30 | 3483 | 5102 | 5251 | 9878 | 11051 | 19506 | 14588 | 10865 |
| 31 | 4632 | 5241 | 6058 | 11264 | 8183 | 12781 | 12680 | 8630 |
| 32 | 4325 | 5908 | 6417 | 10276 | 6916 | 9017 | 12436 | 8234 |
| 33 | 4624 | 4376 | 6451 | 8759 | 5102 | 6395 | 10817 | 6731 |
| 34 | 4281 | 3841 | 6256 | 7187 | 3791 | 4008 | 10013 | 5694 |
| 35 | 3913 | 2637 | 4823 | 5945 | 2717 | 3308 | 9194 | 4647 |
| 36 | 2742 | 1808 | 3900 | 4696 | 1922 | 2572 | 6725 | 4032 |
| 37 | 2705 | 1440 | 3134 | 2743 | 1356 | 1757 | 4337 | 3397 |
| 38 | 2460 | 1372 | 2422 | 2796 | 906 | 1474 | 3548 | 3106 |
| 39 | 1628 | 1025 | 1616 | 2011 | 675 | 1148 | 2294 | 2569 |
| 40 | 1606 | 495 | 1071 | 1137 | 410 | 909 | 1713 | 1627 |
| 41 | 847 | 592 | 893 | 730 | 398 | 938 | 1606 | 1831 |
| 42 | 825 | 549 | 554 | 573 | 193 | 447 | 1192 | 864 |
| 43 | 474 | 414 | 365 | 305 | 152 | 481 | 726 | 797 |
| 44 | 438 | 203 | 233 | 186 | 135 | 142 | 470 | 517 |
| 45 | 300 | 110 | 156 | 41 | 82 | 148 | 297 | 240 |
| 46 | 183 | 131 | 72 | 163 | 32 | 138 | 123 | 293 |
| 47 | 126 | 167 | 106 | 27 | 69 | 64 | 40 | 137 |
| 48 | 195 | 70 | 61 | 68 | 18 | 50 | 81 | 169 |
| 49 | 121 | 77 | 51 | 0 | 16 | 89 | 207 | 40 |
| 50 | 29 | 40 | 14 | 41 | 5 | 0 | 165 | 51 |
| 51 | 18 | 60 | 10 | 0 | 0 | 0 | 61 | 29 |
| 52 | 18 | 24 | 7 | 0 | 13 | 0 | 41 | 68 |
| 53 | 0 | 0 | 12 | 0 | 0 | 24 | 0 | 43 |
| 54 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 171 |
| 55 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 86 |
| 56 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 171 |
| 57 | 18 | 10 | 19 | 0 | 0 | 0 | 0 | 57 |
| 58 | 18 | 0 | 3 | 0 | 0 | 14 | 0 | 86 |
| 59 | 0 | 0 | 8 | 0 | 13 | 0 | 0 | 57 |
| 60 | 18 | 0 | 1 | 0 | 0 | 0 | 0 | 86 |
| 61 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 71 |
| 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 43 |
| 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 29 |
| 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 |
| 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 66 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 67 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 68 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 69 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 14 |
| 70 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 71 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 72 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 74 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 58378 | 67953 | 62236 | 112807 | 108006 | 201588 | 155956 | 127915 |
| Weights | 1496 | 1388 | 1627 | 2391 | 1864 | 3213 | 3411 | 2844 |

Table 7.7.13b. Nephrops in FUs 20-22 Celtic Sea (VIIfgh). Discards-length distributions in 20022009. Irish trawlers.

| Total Discards |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CL mm/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 1 | 0 | 19 | 0 | 0 | 0 |
| 14 | 0 | 19 | 2 | 0 | 38 | 40 | 0 | 0 |
| 15 | 0 | 84 | 3 | 0 | 206 | 0 | 35 | 25 |
| 16 | 0 | 68 | 5 | 35 | 138 | 153 | 70 | 0 |
| 17 | 0 | 171 | 15 | 35 | 243 | 200 | 181 | 178 |
| 18 | 0 | 261 | 16 | 164 | 364 | 772 | 320 | 300 |
| 19 | 0 | 614 | 77 | 265 | 564 | 1784 | 744 | 644 |
| 20 | 0 | 1489 | 49 | 786 | 883 | 3919 | 1372 | 1266 |
| 21 | 1 | 3118 | 94 | 1120 | 1687 | 7572 | 1854 | 1273 |
| 22 | 3 | 4657 | 125 | 1512 | 2993 | 11791 | 2848 | 3018 |
| 23 | 12 | 5158 | 215 | 2213 | 3393 | 15300 | 4324 | 3688 |
| 24 | 48 | 4482 | 358 | 3103 | 4829 | 17669 | 6275 | 5080 |
| 25 | 194 | 4164 | 498 | 3957 | 5468 | 17333 | 9561 | 5596 |
| 26 | 1161 | 4026 | 748 | 5023 | 5758 | 13454 | 9047 | 4877 |
| 27 | 618 | 2926 | 787 | 4766 | 6746 | 10606 | 8600 | 4447 |
| 28 | 476 | 2227 | 751 | 3433 | 5199 | 10847 | 6591 | 2991 |
| 29 | 238 | 1556 | 762 | 2667 | 2630 | 5029 | 4500 | 2335 |
| 30 | 163 | 890 | 708 | 2001 | 1071 | 1752 | 3580 | 1650 |
| 31 | 174 | 511 | 635 | 1051 | 607 | 541 | 2652 | 1150 |
| 32 | 162 | 275 | 421 | 622 | 347 | 151 | 1626 | 749 |
| 33 | 103 | 67 | 304 | 65 | 166 | 17 | 905 | 461 |
| 34 | 61 | 0 | 107 | 1 | 64 | 5 | 617 | 236 |
| 35 | 34 | 0 | 92 | 0 | 0 | 1 | 55 | 68 |
| 36 | 19 | 0 | 9 | 0 | 0 | 0 | 27 | 0 |
| 37 | 10 | 0 | 5 | 0 | 0 | 0 | 0 | 0 |
| 38 | 5 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 39 | 3 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| 40 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 41 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 47 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 48 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 51 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 52 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 53 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 54 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 55 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 56 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 57 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 58 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 59 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 60 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 61 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 66 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 67 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 69 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 71 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 72 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 74 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 3488 | 36762 | 6799 | 32820 | 43413 | 118937 | 65784 | 40032 |
| Weights | 49 | 333 | 99 | 371 | 451 | 1097 | 765 | 426 |

Table 7.7.13c. Nephrops in FUs 20-22 Celtic Sea (VIIfgh). Catches-length distributions in 20022009. Irish trawlers.

| Total catches |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CL mm/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 1 | 0 | 19 | 0 | 0 | 0 |
| 14 | 0 | 19 | 2 | 0 | 38 | 40 | 0 | 0 |
| 15 | 0 | 84 | 3 | 0 | 206 | 0 | 35 | 25 |
| 16 | 0 | 68 | 5 | 35 | 138 | 153 | 70 | 0 |
| 17 | 0 | 171 | 15 | 35 | 243 | 200 | 181 | 178 |
| 18 | 0 | 263 | 21 | 167 | 368 | 772 | 320 | 300 |
| 19 | 33 | 624 | 97 | 282 | 579 | 1813 | 744 | 644 |
| 20 | 146 | 1758 | 89 | 987 | 1116 | 4387 | 1372 | 1393 |
| 21 | 214 | 4186 | 171 | 1782 | 2191 | 8173 | 1882 | 1440 |
| 22 | 656 | 5986 | 259 | 2300 | 4116 | 14151 | 3144 | 3584 |
| 23 | 1352 | 7728 | 460 | 3726 | 5727 | 23320 | 5582 | 5331 |
| 24 | 2291 | 6621 | 838 | 6356 | 9383 | 30802 | 9896 | 9287 |
| 25 | 2439 | 6637 | 1304 | 8645 | 11910 | 35475 | 16542 | 13001 |
| 26 | 3620 | 9841 | 2162 | 11777 | 15876 | 36158 | 21444 | 16402 |
| 27 | 3035 | 7296 | 2839 | 13051 | 19291 | 32827 | 18379 | 16778 |
| 28 | 3935 | 8871 | 3687 | 11944 | 18656 | 35329 | 21329 | 14735 |
| 29 | 3343 | 7130 | 4796 | 11977 | 15151 | 29046 | 18002 | 14954 |
| 30 | 3646 | 5992 | 5959 | 11879 | 12123 | 21258 | 18168 | 12515 |
| 31 | 4806 | 5752 | 6693 | 12315 | 8790 | 13322 | 15332 | 9780 |
| 32 | 4487 | 6183 | 6838 | 10898 | 7263 | 9168 | 14062 | 8983 |
| 33 | 4728 | 4443 | 6756 | 8824 | 5268 | 6412 | 11722 | 7192 |
| 34 | 4343 | 3841 | 6362 | 7188 | 3855 | 4013 | 10630 | 5930 |
| 35 | 3948 | 2637 | 4915 | 5946 | 2717 | 3310 | 9249 | 4715 |
| 36 | 2760 | 1808 | 3909 | 4696 | 1922 | 2573 | 6752 | 4032 |
| 37 | 2715 | 1440 | 3139 | 2743 | 1356 | 1757 | 4337 | 3397 |
| 38 | 2465 | 1372 | 2425 | 2796 | 906 | 1474 | 3548 | 3106 |
| 39 | 1631 | 1025 | 1618 | 2011 | 675 | 1148 | 2294 | 2569 |
| 40 | 1608 | 495 | 1072 | 1137 | 410 | 909 | 1713 | 1627 |
| 41 | 848 | 592 | 894 | 730 | 398 | 938 | 1606 | 1831 |
| 42 | 826 | 549 | 554 | 573 | 193 | 447 | 1192 | 864 |
| 43 | 475 | 414 | 366 | 305 | 152 | 481 | 726 | 797 |
| 44 | 438 | 203 | 234 | 186 | 135 | 142 | 470 | 517 |
| 45 | 300 | 110 | 156 | 41 | 82 | 148 | 297 | 240 |
| 46 | 183 | 131 | 72 | 163 | 32 | 138 | 123 | 293 |
| 47 | 126 | 167 | 107 | 27 | 69 | 64 | 40 | 137 |
| 48 | 195 | 70 | 61 | 68 | 18 | 50 | 81 | 169 |
| 49 | 121 | 77 | 51 | 0 | 16 | 89 | 207 | 40 |
| 50 | 29 | 40 | 14 | 41 | 5 | 0 | 165 | 51 |
| 51 | 18 | 60 | 10 | 0 | 0 | 0 | 61 | 29 |
| 52 | 18 | 24 | 7 | 0 | 13 | 0 | 41 | 68 |
| 53 | 0 | 0 | 12 | 0 | 0 | 24 | 0 | 43 |
| 54 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 171 |
| 55 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 86 |
| 56 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 171 |
| 57 | 18 | 10 | 19 | 0 | 0 | 0 | 0 | 57 |
| 58 | 18 | 0 | 3 | 0 | 0 | 14 | 0 | 86 |
| 59 | 0 | 0 | 8 | 0 | 13 | 0 | 0 | 57 |
| 60 | 18 | 0 | 1 | 0 | 0 | 0 | 0 | 86 |
| 61 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 71 |
| 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 43 |
| 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 29 |
| 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 |
| 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 66 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 67 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 68 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 69 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 14 |
| 70 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 71 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 72 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 74 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 61866 | 104715 | 69034 | 145627 | 151419 | 320525 | 221740 | 167947 |
| Weights | 1545 | 1721 | 1727 | 2762 | 2315 | 4311 | 4176 | 3271 |

Table 7.7.13d. Nephrops in FUs 20-22 Celtic Sea (VIIfgh). Removals-length distributions in 20022009. Irish trawlers.

| Removals=Landings+dead catches (discard survival rate : 25\%) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CL mm/Year | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 1 | 0 | 14 | 0 | 0 | 0 |
| 14 | 0 | 14 | 2 | 0 | 29 | 30 | 0 | 0 |
| 15 | 0 | 63 | 2 | 0 | 155 | 0 | 26 | 19 |
| 16 | 0 | 51 | 4 | 26 | 104 | 115 | 53 | 0 |
| 17 | 0 | 128 | 12 | 26 | 182 | 150 | 136 | 134 |
| 18 | 0 | 197 | 18 | 126 | 277 | 579 | 240 | 225 |
| 19 | 33 | 471 | 77 | 216 | 438 | 1367 | 558 | 483 |
| 20 | 145 | 1386 | 77 | 790 | 896 | 3407 | 1029 | 1076 |
| 21 | 214 | 3407 | 148 | 1502 | 1769 | 6280 | 1419 | 1122 |
| 22 | 655 | 4822 | 228 | 1921 | 3368 | 11203 | 2432 | 2830 |
| 23 | 1349 | 6438 | 406 | 3172 | 4879 | 19495 | 4501 | 4409 |
| 24 | 2279 | 5500 | 748 | 5580 | 8176 | 26385 | 8327 | 8017 |
| 25 | 2390 | 5596 | 1179 | 7656 | 10543 | 31142 | 14152 | 11602 |
| 26 | 3330 | 8834 | 1975 | 10521 | 14436 | 32795 | 19183 | 15183 |
| 27 | 2881 | 6564 | 2642 | 11859 | 17604 | 30176 | 16229 | 15666 |
| 28 | 3816 | 8314 | 3499 | 11085 | 17357 | 32618 | 19681 | 13988 |
| 29 | 3283 | 6741 | 4606 | 11310 | 14493 | 27789 | 16877 | 14370 |
| 30 | 3605 | 5769 | 5782 | 11379 | 11855 | 20820 | 17273 | 12102 |
| 31 | 4763 | 5625 | 6534 | 12052 | 8638 | 13187 | 14669 | 9492 |
| 32 | 4446 | 6114 | 6733 | 10743 | 7176 | 9131 | 13656 | 8796 |
| 33 | 4702 | 4426 | 6680 | 8808 | 5226 | 6408 | 11496 | 7077 |
| 34 | 4327 | 3841 | 6336 | 7187 | 3839 | 4012 | 10476 | 5871 |
| 35 | 3939 | 2637 | 4892 | 5946 | 2717 | 3309 | 9236 | 4698 |
| 36 | 2756 | 1808 | 3906 | 4696 | 1922 | 2573 | 6745 | 4032 |
| 37 | 2712 | 1440 | 3138 | 2743 | 1356 | 1757 | 4337 | 3397 |
| 38 | 2464 | 1372 | 2424 | 2796 | 906 | 1474 | 3548 | 3106 |
| 39 | 1630 | 1025 | 1617 | 2011 | 675 | 1148 | 2294 | 2569 |
| 40 | 1607 | 495 | 1072 | 1137 | 410 | 909 | 1713 | 1627 |
| 41 | 847 | 592 | 894 | 730 | 398 | 938 | 1606 | 1831 |
| 42 | 825 | 549 | 554 | 573 | 193 | 447 | 1192 | 864 |
| 43 | 475 | 414 | 366 | 305 | 152 | 481 | 726 | 797 |
| 44 | 438 | 203 | 233 | 186 | 135 | 142 | 470 | 517 |
| 45 | 300 | 110 | 156 | 41 | 82 | 148 | 297 | 240 |
| 46 | 183 | 131 | 72 | 163 | 32 | 138 | 123 | 293 |
| 47 | 126 | 167 | 107 | 27 | 69 | 64 | 40 | 137 |
| 48 | 195 | 70 | 61 | 68 | 18 | 50 | 81 | 169 |
| 49 | 121 | 77 | 51 | 0 | 16 | 89 | 207 | 40 |
| 50 | 29 | 40 | 14 | 41 | 5 | 0 | 165 | 51 |
| 51 | 18 | 60 | 10 | 0 | 0 | 0 | 61 | 29 |
| 52 | 18 | 24 | 7 | 0 | 13 | 0 | 41 | 68 |
| 53 | 0 | 0 | 12 | 0 | 0 | 24 | 0 | 43 |
| 54 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 171 |
| 55 | 18 | 0 | 6 | 0 | 0 | 0 | 0 | 86 |
| 56 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 171 |
| 57 | 18 | 10 | 19 | 0 | 0 | 0 | 0 | 57 |
| 58 | 18 | 0 | 3 | 0 | 0 | 14 | 0 | 86 |
| 59 | 0 | 0 | 8 | 0 | 13 | 0 | 0 | 57 |
| 60 | 18 | 0 | 1 | 0 | 0 | 0 | 0 | 86 |
| 61 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 71 |
| 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 43 |
| 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 29 |
| 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 |
| 65 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 66 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 67 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 68 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 69 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 14 |
| 70 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 |
| 71 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 72 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 74 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 60994 | 95525 | 67335 | 137422 | 140565 | 290791 | 205294 | 157939 |
| Weights | 1533 | 1638 | 1702 | 2669 | 2202 | 4036 | 3984 | 3164 |

Table 7.7.14. Nephrops in VIIfgh. Mean sizes (carapace length, CL in mm) of French and Irish landings. For 2008 and 2009, French values are calculated (1) including the samples involving in tailed individuals and (2) using the previous method (no sampling of tails; the total tailed proportion was apportioned in the smallest category of entire Nephrops at auction).

| Year | French sampling |  |  | Irish sampling |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Total | Males | Females | Total |
| 1987 | 38.8 | 35.1 | 38.1 |  |  |  |
| 1988 | 35.7 | 34.7 | 35.6 |  |  |  |
| 1989 | 38.9 | 36.0 | 38.5 |  |  |  |
| 1990 | 39.7 | 35.4 | 39.0 |  |  |  |
| 1991 | 38.2 | 34.1 | 37.5 |  |  |  |
| 1992 | 37.6 | 34.9 | 37.3 |  |  |  |
| 1993 | 40.0 | 36.6 | 39.6 |  |  |  |
| 1994 | 39.7 | 37.1 | 39.3 |  |  |  |
| 1995 | 39.9 | 36.1 | 39.4 |  |  |  |
| 1996 | 39.5 | 36.8 | 39.2 |  |  |  |
| 1997 | 39.9 | 37.4 | 39.8 |  |  |  |
| 1998 | 39.9 | 36.4 | 39.5 |  |  |  |
| 1999 | 40.1 | 36.9 | 39.6 |  |  |  |
| 2000 | 42.0 | 39.2 | 41.4 |  |  |  |
| 2001 | 38.8 | 39.1 | 38.9 |  |  |  |
| 2002 | 40.9 | 39.7 | 40.8 | 33.0 | 31.1 | 32.2 |
| 2003 | 41.5 | 39.8 | 41.4 | 31.1 | 29.1 | 30.2 |
| 2004 | 41.6 | 39.8 | 41.5 | 33.5 | 32.3 | 32.9 |
| 2005 | 43.1 | 40.3 | 42.8 | 30.9 | 30.8 | 30.9 |
| 2006 | 41.6 | 39.5 | 41.1 | 29.7 | 28.6 | 29.2 |
| 2007 | 40.7 | 38.7 | 40.4 | 29.3 | 27.3 | 28.5 |
| 2008 | 37.6 | 34.7 | 37.2 | 32.0 | 29.7 | 31.1 |
|  | 40.1 | 39.6 | 40.1 |  |  |  |
| 2009 | 39.0 | 34.5 | 38.6 | 31.8 | 28.8 | 30.8 |
|  | 41.0 | 40.1 | 41.0 |  |  |  |

Note: French values for the period 1999-2007 are provisional (they will be revised after validation of the revision for DLF including tailed Nephrops).

Table 7.7.15. Nephrops in VIIfgh. French (year 1997) and Irish (years 2003, 2008 and 2009) programs of discard sampling onboard. Length distribution of landings (L) and discards ( $\mathrm{D)} \mathrm{by} \mathrm{sex}\left(10^{3}\right)$. The reported size is the carapace length (CL, in mm ). Conversion of CL to TS (total size) is done by multiplication by 3.3 .

|  | French sampling (year 1997) |  |  |  |  |  | Irish sampling (year 2003) |  |  |  |  |  | Irish sampling (year 2008) |  |  |  |  |  | Irish sampling (year 2009) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | males |  | females |  | Total |  | males |  | females |  | Total |  | males |  | females |  | Total |  | males |  | females |  | Total |  |
| CL | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D |
| 14 |  |  |  |  |  |  |  |  |  | 19 |  | 19 |  |  |  |  |  |  |  |  |  |  |  |  |
| 15 |  |  |  |  |  |  |  | 10 |  | 74 |  | 84 |  |  |  | 35 |  | 35 |  | 25 |  |  |  | 25 |
| 16 |  |  |  | 1 |  | 1 |  | 10 |  | 58 |  | 68 |  | 35 |  | 35 |  | 70 |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  | 30 |  | 141 |  | 171 |  |  |  | 181 |  | 181 |  | 42 |  | 136 |  | 178 |
| 18 |  |  |  | 1 |  | 1 | 1 | 134 | 1 | 127 | 2 | 261 |  | 83 |  | 237 |  | 320 |  | 77 |  | 223 |  | 300 |
| 19 |  | 1 |  |  |  | 1 | 3 | 242 | 7 | 372 | 10 | 614 |  | 166 |  | 578 |  | 744 |  | 239 |  | 405 |  | 644 |
| 20 |  | 1 |  | 12 |  | 13 | 53 | 452 | 215 | 1038 | 269 | 1489 |  | 370 |  | 1002 |  | 1372 |  | 247 | 127 | 1019 | 127 | 1266 |
| 21 |  |  |  | 10 |  | 10 | 164 | 902 | 904 | 2216 | 1068 | 3118 |  | 988 | 28 | 866 | 28 | 1854 | 111 | 679 | 56 | 594 | 167 | 1273 |
| 22 |  | 187 |  | 294 |  | 481 | 472 | 1963 | 858 | 2693 | 1330 | 4657 | 98 | 1127 | 198 | 1721 | 296 | 2848 | 220 | 1182 | 346 | 1836 | 566 | 3018 |
| 23 |  | 630 |  | 1150 |  | 1780 | 1469 | 2503 | 1101 | 2655 | 2570 | 5158 | 195 | 1431 | 1063 | 2893 | 1258 | 4324 | 756 | 1610 | 887 | 2078 | 1643 | 3688 |
| 24 |  | 874 |  | 1172 |  | 2046 | 1251 | 2392 | 888 | 2091 | 2139 | 4482 | 1491 | 2022 | 2130 | 4253 | 3621 | 6275 | 2015 | 2235 | 2192 | 2845 | 4207 | 5080 |
| 25 |  | 1428 |  | 2490 |  | 3918 | 1209 | 2056 | 1264 | 2109 | 2473 | 4164 | 3058 | 2931 | 3923 | 6630 | 6981 | 9561 | 4121 | 2814 | 3284 | 2782 | 7405 | 5596 |
| 26 |  | 1439 |  | 1889 |  | 3328 | 3132 | 1631 | 2683 | 2396 | 5815 | 4026 | 5878 | 2971 | 6519 | 6076 | 12397 | 9047 | 5814 | 2316 | 5711 | 2561 | 11525 | 4877 |
| 27 | 15 | 4695 |  | 7332 | 15 | 12027 | 1978 | 1304 | 2392 | 1622 | 4370 | 2926 | 4798 | 3416 | 4981 | 5184 | 9779 | 8600 | 6595 | 2292 | 5735 | 2155 | 12331 | 4447 |
| 28 | 28 | 4399 |  | 6888 | 28 | 11287 | 3591 | 1030 | 3053 | 1196 | 6644 | 2227 | 8319 | 3258 | 6419 | 3333 | 14738 | 6591 | 6508 | 1644 | 5236 | 1347 | 11744 | 2991 |
| 29 | 45 | 3521 |  | 5089 | 45 | 8610 | 2568 | 723 | 3006 | 833 | 5574 | 1556 | 8292 | 2362 | 5209 | 2138 | 13502 | 4500 | 7532 | 1311 | 5087 | 1024 | 12619 | 2335 |
| 30 | 218 | 6863 | 19 | 9305 | 236 | 16167 | 2327 | 433 | 2775 | 457 | 5102 | 890 | 9274 | 1926 | 5314 | 1654 | 14588 | 3580 | 6985 | 1076 | 3879 | 574 | 10865 | 1650 |
| 31 | 521 | 3140 | 21 | 4821 | 542 | 7960 | 2977 | 300 | 2265 | 211 | 5241 | 511 | 7186 | 1431 | 5495 | 1221 | 12680 | 2652 | 5539 | 751 | 3091 | 399 | 8630 | 1150 |
| 32 | 1155 | 4842 | 65 | 6535 | 1220 | 11377 | 3570 | 166 | 2338 | 109 | 5908 | 275 | 7137 | 914 | 5299 | 712 | 12436 | 1626 | 5748 | 580 | 2486 | 169 | 8234 | 749 |


| CL | French sampling (year 1997) |  |  |  |  |  | Irish sampling (year 2003) |  |  |  |  |  | Irish sampling (year 2008) |  |  |  |  |  | Irish sampling (year 2009) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | males |  | females |  | Total |  | males |  | females |  | Total |  | males |  | females |  | Total |  | males |  | females |  | Total |  |
|  | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D | L | D |
| 33 | 1984 | 3885 | 160 | 5140 | 2144 | 9025 | 2313 | 57 | 2063 | 9 | 4376 | 67 | 7181 | 585 | 3636 | 320 | 10817 | 905 | 4680 | 388 | 2051 | 73 | 6731 | 461 |
| 34 | 2035 | 1360 | 152 | 1384 | 2186 | 2744 | 2371 |  | 1470 |  | 3841 |  | 7008 | 332 | 3005 | 285 | 10013 | 617 | 4353 | 220 | 1341 | 16 | 5694 | 236 |
| 35 | 3251 | 1385 | 357 | 1254 | 3608 | 2639 | 1468 |  | 1168 |  | 2637 |  | 6570 | 55 | 2624 |  | 9194 | 55 | 3721 | 68 | 926 |  | 4647 | 68 |
| 36 | 3409 | 570 | 418 | 950 | 3827 | 1520 | 1108 |  | 700 |  | 1808 |  | 5201 | 27 | 1524 |  | 6725 | 27 | 3236 |  | 797 |  | 4032 |  |
| 37 | 3799 | 410 | 464 | 333 | 4262 | 743 | 1056 |  | 384 |  | 1440 |  | 3430 |  | 906 |  | 4337 |  | 2864 |  | 533 |  | 3397 |  |
| 38 | 4138 | 205 | 666 | 189 | 4804 | 394 | 1140 |  | 232 |  | 1372 |  | 2993 |  | 556 |  | 3548 |  | 2785 |  | 321 |  | 3106 |  |
| 39 | 3395 | 72 | 224 | 85 | 3619 | 157 | 891 |  | 134 |  | 1025 |  | 1928 |  | 366 |  | 2294 |  | 2334 |  | 235 |  | 2569 |  |
| 40 | 4713 | 120 | 205 | 64 | 4918 | 184 | 404 |  | 91 |  | 495 |  | 1526 |  | 187 |  | 1713 |  | 1411 |  | 216 |  | 1627 |  |
| 41 | 2861 | 33 | 202 | 41 | 3062 | 74 | 572 |  | 20 |  | 592 |  | 1459 |  | 148 |  | 1606 |  | 1667 |  | 163 |  | 1831 |  |
| 42 | 3367 | 43 | 47 | 34 | 3414 | 77 | 492 |  | 57 |  | 549 |  | 1114 |  | 78 |  | 1192 |  | 827 |  | 37 |  | 864 |  |
| 43 | 2678 | 25 | 47 |  | 2725 | 25 | 386 |  | 29 |  | 414 |  | 650 |  | 76 |  | 726 |  | 766 |  | 32 |  | 797 |  |
| 44 | 1787 | 8 | 63 |  | 1849 | 8 | 155 |  | 48 |  | 203 |  | 431 |  | 40 |  | 470 |  | 503 |  | 14 |  | 517 |  |
| 45 | 2236 | 7 | 52 | 2 | 2288 | 9 | 110 |  |  |  | 110 |  | 297 |  |  |  | 297 |  | 226 |  | 15 |  | 240 |  |
| 46 | 1428 | 1 |  |  | 1428 | 1 | 131 |  |  |  | 131 |  | 123 |  |  |  | 123 |  | 270 |  | 23 |  | 293 |  |
| 47 | 1021 |  |  |  | 1021 |  | 167 |  |  |  | 167 |  | 40 |  |  |  | 40 |  | 137 |  |  |  | 137 |  |
| 48 | 954 | 2 | 16 |  | 970 | 2 | 70 |  |  |  | 70 |  | 81 |  |  |  | 81 |  | 169 |  |  |  | 169 |  |
| 49 | 603 |  |  |  | 603 |  | 77 |  |  |  | 77 |  | 207 |  |  |  | 207 |  | 40 |  |  |  | 40 |  |
| 50 | 733 | 1 |  |  | 733 | 1 | 40 |  |  |  | 40 |  | 165 |  |  |  | 165 |  | 51 |  |  |  | 51 |  |
| 51 | 353 |  |  |  | 353 |  | 60 |  |  |  | 60 |  | 61 |  |  |  | 61 |  | 29 |  |  |  | 29 |  |
| 52 | 372 |  |  |  | 372 |  | 24 |  |  |  | 24 |  | 41 |  |  |  | 41 |  | 57 |  | 11 |  | 68 |  |
| 53 | 286 | 3 |  |  | 286 | 3 |  |  |  |  |  |  |  |  |  |  |  |  | 43 |  |  |  | 43 |  |
| 54 | 198 |  |  |  | 198 |  |  |  |  |  |  |  |  |  |  |  |  |  | 171 |  |  |  | 171 |  |



Table 7.7.16. Nephrops in the Celtic Sea (FU20-22). Production by rectangle for French and Irish trawlers. The total by rectangle and the \% involve in years 1999-2008 for French fleet and in years 2003-2009 for Irish fleet. Rectangles associated with the "Smalls" ground are highlighted in grey.

|  | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | Total | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| French trawlers |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 28E1 | 78 | 75 | 127 | 207 | 246 | 164 | 191 | 212 | 375 | 362 |  | 2038 | 8\% |
| 28E2 | 146 | 350 | 331 | 287 | 363 | 259 | 296 | 214 | 189 | 252 |  | 2687 | 10\% |
| 29E1 | 105 | 182 | 302 | 535 | 653 | 353 | 277 | 258 | 398 | 354 |  | 3417 | 13\% |
| 29E2 | 129 | 287 | 205 | 204 | 249 | 261 | 371 | 423 | 240 | 223 |  | 2593 | 10\% |
| 30 E 1 | 121 | 170 | 205 | 437 | 374 | 205 | 179 | 104 | 106 | 146 |  | 2048 | 8\% |
| 30E2 | 293 | 424 | 434 | 741 | 806 | 781 | 577 | 773 | 437 | 661 |  | 5928 | 23\% |
| 31 E 3 | 847 | 1016 | 763 | 489 | 679 | 396 | 423 | 249 | 193 | 230 |  | 5285 | 20\% |
| Irish trawlers |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 28E1 |  |  |  | 0 | 6 | 4 | 10 | 2 | 10 | 18 | 64 | 114 | 1\% |
| 28E2 | 0 | 3 | 1 | 1 | 2 | 23 | 15 | 6 | 2 | 6 | 72 | 131 | 1\% |
| 29E1 | 13 | 18 | 0 | 9 | 34 | 38 | 105 | 91 | 194 | 374 | 476 | 1,352 | 6\% |
| 29E2 | 1 | 2 |  |  | 1 | 11 | 19 | 24 | 31 | 23 | 67 | 179 | 1\% |
| 30 E 1 | 5 | 11 | 28 | 39 | 62 | 104 | 133 | 141 | 154 | 292 | 297 | 1,265 | 6\% |
| 30 E 2 | 3 | 4 | 3 | 2 | 5 | 36 | 52 | 99 | 69 | 147 | 151 | 570 | 3\% |
| 30E3 | 13 | 9 | 0 | 5 | 2 | 27 | 55 | 39 | 40 | 15 | 16 | 221 | 1\% |
| 31 E 2 | 39 | 45 | 53 | 49 | 37 | 56 | 68 | 49 | 101 | 61 | 59 | 615 | 3\% |
| 31 E 3 | 544 | 1165 | 1628 | 1103 | 941 | 1101 | 1571 | 1168 | 2392 | 2257 | 1549 | 15,419 | 69\% |
| 31E4 | 24 | 21 | 142 | 130 | 115 | 17 | 129 | 85 | 96 | 61 | 40 | 859 | 4\% |
| 32E3 | 86 | 195 | 222 | 130 | 185 | 211 | 231 | 145 | 126 | 156 | 53 | 1,740 | 8\% |

Table 7.7.17. Division VIIfgh. Nephrops effort and lpue data by country. The French data are calculated for otter trawlers getting at least $10 \%$ of their landings by targeting this species. The Irish data are linked to otter trawl vessels where $>30 \%$ of monthly landings in live weight were Nephrops.

|  | Effort <br> (Effective hours fishing) |  | Ipue (kg/h) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | France | Rep. of Ireland | France |  |  | Rep. of |
|  | Otter | Otter | Total otter | Single <br> Otter ${ }^{13}$ | Twin otter 13 | Ireland <br> Otter |
| Year |  |  |  |  |  |  |
| 1983 | 231440 |  | 14.2 | 14.2 |  |  |
| 1984 | 204600 |  | 15.8 | 15.8 |  |  |
| 1985 | 202830 |  | 16.0 | 16.0 |  |  |
| 1986 | 162510 |  | 14.9 | 14.9 |  |  |
| 1987 | 189580 |  | 15.2 | 15.2 |  |  |
| 1988 | 170840 |  | 16.4 | 16.4 |  |  |
| 1989 | 179060 |  | 16.8 | 16.8 |  |  |
| 1990 | 229470 |  | 15.6 | 15.6 |  |  |
| 1991 | 224710 |  | 11.3 | 11.3 |  |  |
| 1992 | 276450 |  | 11.7 | 11.7 |  |  |
| 1993 | 268410 |  | 13.2 | 13.2 |  |  |
| 1994 | 258490 |  | 13.5 | 13.5 |  |  |
| 1995 | 239240 | 26681 | 14.6 | 14.6 |  | 46.9 |
| 1996 | 220120 | 20579 | 14.2 | 14.2 | 14.2 | 50.0 |
| 1997 | 187180 | 23255 | 12.6 | 12.5 | 14.4 | 49.2 |
| 1998 | 155340 | 25380 | 13.0 | 12.9 | 14.9 | 53.1 |
| 1999 | 150770 | 15491 | 10.9 | 10.2 | 10.0 | 41.5 |
| 2000 | 194150 | 28267 | 13.8 | 11.5 | 11.4 | 47.8 |
| 2001 | 170320 | 36205 | 14.6 | 11.4 | 13.3 | 54.6 |
| 2002 | 165670 | 29990 | 18.7 | 15.4 | 16.7 | 44.3 |
| 2003 | 191600 | 28532 | 18.2 | 16.3 | 15.0 | 33.9 |
| 2004 | 152700 | 31309 | 15.8 | 13.5 | 12.9 | $32.8{ }^{4}$ |
| 2005 | 146880 | 51031 | 16.0 | 13.0 | 13.2 | $41.3^{4}$ |
| 2006 | 136650 | 45383 | 16.3 | 14.4 | 12.8 | $34.9{ }^{4}$ |
| 2007 | 101980 | 59899 | 18.5 | 15.9 | 14.3 | $48.1^{4}$ |
| 2008 | 99789 | 59875 | 22.6 | 18.4 | 16.4 | $53.8{ }^{4}$ |
| 2009 | na | $55454{ }^{4}$ | $n a^{5}$ | $n a^{5}$ | $n a^{5}$ | 48.2 |

${ }^{1}$ The single and twin otter French lpue can be compared with the total otter indices until 1999 when the definition of the fishing effort of trawlers was changed (see note 2 ).
${ }^{2}$ For the period 1999-2008, the French statistics differentiate fishing effort calculated on the basis of the "number of fishing hours" from that deduced from the "number of use of a fishing gear".
${ }^{3}$ Information for single and twin trawl lpue involve in the total fishing fleet whereas aggregated indices are calculated for the otter trawlers getting at least $10 \%$ of their landings by targeting this species.
${ }^{4}$ Revised data (WGCSE2010).
${ }^{5}$ Not available.


Figure 7.7.1. Nephrops in VIIfgh. Evolution of nominal landings $(\mathbf{t})$.


Figure 7.7.2. Nephrops in FU 20-22 Celtic Sea (VIIfgh) landings of French trawlers (1987-2009) and of Irish trawlers (2002-2009). French landings since 1999 are presented by two ways: (1) Lines: previous method (tails not sampled and systematically apportioned in the smallest category of entire Nephrops at auction).(2) Bars: tails are included (years 1999-2007: simulation; since 2008: sampled data).


Figure 7.7.3. Nephrops of the Celtic Sea (VIIfgh, FU20-22). Years 1999-2009. Monthly percentages of tailed individuals in the French landings (after conversion to total weight).


Figure 7.7.4. Nephrops of the Celtic Sea (VIIfgh, FU20-22). French landings for 2008. Length distributions (1) including the data on tails and (2) using the previous method (no sampling of tails; the total tailed proportion was apportioned in the smallest category of entire Nephrops at auction).


Figure 7.7.5. Nephrops of the Celtic Sea (VIIfgh, FU20-22). French landings for 2009 by sex. Length distributions (1) including the data on tails and (2) using the previous method (no sampling of tails; the total tailed proportion was apportioned in the smallest category of entire Nephrops at auction).



Figure 7.7.6. Nephrops in FU 20-22 Celtic Sea (VIIfgh). Years with complete set of discard samples: French data (1997), Irish data (2003, 2008 and 2009). Landings in white, discards in black.


Figure 7.7.7. Nephrops in FU 20-22 Celtic Sea (VIIfgh). Catches (landings in white and discards in black) of the Irish fleet. Length distributions in 2002-2009.

| Ground | Year | Number of stations | Mean Density <br> (No./M2) | Domain Area (m2) | Geostatistical <br> abundance estimate <br> (million burrows) | CV on Burrow <br> estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2006 | 100 | 0.62 | 2847 | 1914 | $3 \%$ |
|  | 2007 | 107 | 0.46 | 2915 | 1402 | $6 \%$ |
| Smalls | 2008 | 76 | 0.47 | 2698 | 1448 | $6 \%$ |



Figure 7.7.8. Nephrops in the Celtic Sea (FU 20-22).Summary of geostatistics results 2006-2009 of the Irish UWTV survey carried out on the Smalls ground (ICES statistical rectangle 31E3) and contour plots of burrow densities.



Figure 7.7.9. Nephrops in VIIfgh. Lpue and fishing effort series for French and Irish fleet. The cpue indices are calculated by including discard sampling onboard. For French data, calculations of discards based on the derivation method (performed by WGSSDS 2006-2008; see Stock Annex) were not re-validated after revision of landings by including tailed individuals, thus, they are not presented.

### 7.8 Nephrops in Divisions VIIjg (South and SW Ireland, FU19)

Type of assessment in 2010
ICES is providing new advice for this stock this year so the Report consists of an update to available data.

ICES advice applicable to 2009 and 2010
The current fishery appears sustainable. Therefore, ICES recommends that Nephrops fisheries should not be allowed to increase relative to 2007. This corresponds to landings of no more than 800 tonnes for the Ireland SW and SE Coast (FU19).

### 7.8.1 General

## Stock description and management units

In FU19 Nephrops are caught on a large number of spatially discrete small inshore grounds and on some larger grounds further offshore Figure 7.8.1. Of these the 'Galley ground', around the Kinsale Gas Rigs and south of Cork appear to be the most important.


A map of the spatial distribution of FU19 is given in the FU includes Nephrops within the following ICES statistical rectangles; 31-33 D9-E0; 31E1; 32E1-E2; 33E2-E3.

### 7.8.2 Fishery description

The number of Irish vessels reporting landings in this area has increased from 28 in 2000 to 82 in 2009. Of these, only 14 reported landings in excess of $10 t$ and these 14 vessels accounted for $64 \%$ of the total landings. Fleet segmentation data shows that
the Nephrops métiers in this area also have important catches of megrim and monkfish. There are also some catches of hake and the offshore parts of FU19 which is an important nursery area for juvenile hake. The Irish fleet fishing Nephrops in FU19 was described in detail in the 2001 WG Report (ICES, 2001a). The minimum mesh size in use is 70 mm , with the average being 80 mm . French trawlers harvesting Nephrops on this area fish also in the Celtic Sea (FU20) and switch to the FU19 according to meteorological conditions. They have used mesh size 100 mm for codend since January 2000 (in order to not be constrained by bycatch composition) and they apply MLS of 11.5 cm (i.e. $35 \mathrm{~mm} C L$ ) adopted by French Producers' Organizations larger than the European one ( 8.5 cm i.e. 25 mm CL). However, the increasing proportion of tailed individuals in French landings (as for FU20) may shift DLF for Nephrops to smaller sizes compared with previous years. In 2009, 20 French trawlers reported landings from FU19, but only three exceeded 5 tonnes. 24 French vessels were recorded in 2008, 31 in 2007, 30 in 2006 and 35 in 2005.

### 7.8.3 Data

The sampling level for the species is given in Table 2.1.

### 7.8.4 Commercial catches and discards

Landings data for FU19 are summarised in Table 7.8.1. The Republic of Ireland, France and the UK report landings for FU19. The Republic of Ireland landings have fluctuated considerably throughout the time-series, with a marked dip in 1994 (Figure 7.8.2). The highest landings in the time-series were observed in 2002-2004 ( $>1000 \mathrm{t}$ ). Landings in 2005 and 2006 have been below average for the series. In 2009 landings decreased by approx. $1 \%$ for the Irish fleet but were above the series average. Landings by the French fleet have fluctuated with a declining trend throughout the time-series from the highest value in 1989 of 245 t to 55 t in 2009. Landings from the UK are minor.

Effort and lpue data are available for the Irish Nephrops directed fleet in FU19 from 1995-2009 (Table 7.8.2, Figure 7.8.4.2). The effort increases substantially in 2002 this is in part due to the inclusion of smaller vessels $(10-18 \mathrm{~m})$ in the dataset. These vessels did not record logbook operations prior 2002. The lpue and effort-series is based on the same criteria for FU16 and 17 and will be contingent on the accuracy of landings data reported in logbooks. The lpues have fluctuated between $15-30 \mathrm{~kg} / \mathrm{hr}$ with a slightly declining trend. The lpues are lower than that of other FUs reflecting the smaller size of the vessels and generally more mixed nature of this fishery.

For FU18 landings information from 1993 was available to the WG only. The Republic of Ireland has taken $100 \%$ of the landings for the last seven years. The highest reported landings were in 1994 with 124 t landings in recent years have been minor (10 t in 2009).

### 7.8.5 Biological sampling

Length frequency data of the landings were collected on an irregular basis in the years 1996 to 1997, 1999 and 2002 to 2006. Spatial and temporal coverage is also problematic with landings from FU19 coming from several discrete grounds. In 2005 length frequency data are only available for quarters 2 and 3 . The length frequencies for the remaining quarters have been derived by raising those length frequencies observed to the quarter 1 and 4 landings figures.

The dataseries of the mean sizes of Nephrops in the landings of Irish trawlers is too short and inconsistent to draw definite conclusions (Table 7.8.3 Figure 7.8.4). The
mean size of males varied between 29 and 41 mm CL, and for females between 26 and 40 mm CL. There is a decrease in mean size for males and females in 2009. However, the dataseries is too short to provide useful information on the state of the stock.

It should be noted that due to the change in sampling methodology from 2001 onwards the profile of the length frequencies has changed as a result of inclusion of smaller individuals from the discard component.

### 7.8.6 Information from surveys

The UK March groundfish survey has been carried between 1984 and 2004. This survey was examined in 2006 and there is a slight indication of a decline in mean sizes of Nephrops compared with those observed in the late 1980s. In 2006 some UWTV stations were carried out within FU19 as part of the Celtic Sea UWTV survey (which mainly targets FU20-22). The heterogeneous distribution of Nephrops and sediment in FU19 will make accurate UWTV survey abundance estimate difficult to obtain on a regular basis.

### 7.8.7 Assessment

A much improved and longer historical time-series of data is needed to carry out analytical assessment of this stock. Although sampling of this stock is required under the EU data collection regulation it is difficult to obtain precise length frequency data at the spatial resolution required to assess Nephrops in such a heterogeneous area where several small discrete fisheries occur. Future assessments would benefit from a higher spatial resolution of landings and effort data (possibly from VMS as in Figure 7.8.1). Fishery independent methods such as UWTV surveys may also be useful for this FU in the future.

### 7.8.8 Management considerations

The time-series of lpue data based on logbook data for FU19 is short and variable but is without an obvious trend. Reported landings in 2009 have been around $4 \%$ above series average.

Nephrops fisheries in this area are fairly mixed also catching megrim, anglerfish and other demersal species. There are also some catches of hake, and the offshore parts of the area. The Nephrops grounds in FU19 coincide with an important nursery area for juvenile hake and anglerfish among other species (ICES, 2009).

ICES has repeatedly advised that management should be at a smaller scale than the ICES Subarea VII. Management at the Functional Unit level could provide the controls to ensure that catch opportunities and effort are at the same scale as the resource. A time-series of landings by all FUs in ICES Subarea VII together with the overall TAC is shown in Table 7.8.4. (Note that national quotas for Ireland and the UK are restrictive in most of the recent years).

### 7.8.9 References

ICES. 2009. Review of the Biologically Sensitive Area/Irish Box http://www.ices.dk/committe/acom/comwork/report/2009/Special\ Requests/EC\ Iris h\%20box.pdf.

Table 7.8.1. Nephrops in FU18 and FU19 (NW, SW and SE Ireland). Landings in tonnes by country and Functional Unit.

| Year | FU 18 |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Rep. of <br> Ireland | UK | Total |  |
| 1989 |  |  |  |  |
| 1990 |  |  |  |  |
| 1991 |  |  | 10 |  |
| 1992 | 9 | 1 | 126 |  |
| 1993 | 124 | 2 | 26 |  |
| 1994 | 24 | 2 | 46 |  |
| 1995 | 46 | 1 | 15 |  |
| 1996 | 13 | 2 | 78 |  |
| 1997 | 77 | 1 | 16 |  |
| 1998 | 15 | 0 | 9 |  |
| 1999 | 9 | 0 | 2 |  |
| 2000 | 2 | 0 | 14 |  |
| 2001 | 14 | 0 | 16 |  |
| 2002 | 16 | 0 | 22 |  |
| 2003 | 22 | 0 | 15 |  |
| 2004 | 15 | 0 | 14 |  |
| 2005 | 14 | 0 | 3 |  |
| 2006 | 3 | 0 | 1 |  |
| 2007 | 1 | 0 | 10 |  |
| 2008 | 10 | 0 |  |  |
| 2009 |  |  |  |  |


| FU 19 |  |  |  |
| :---: | :---: | :---: | :---: |
| France | Rep. of <br> Ireland | UK | Total |
| 245 | 652 | 2 | 899 |
| 181 | 569 | 4 | 754 |
| 212 | 860 | 5 | 1077 |
| 233 | 640 | 15 | 888 |
| 229 | 672 | 4 | 905 |
| 216 | 153 | 21 | 390 |
| 175 | 507 | 12 | 695 |
| 145 | 736 | 7 | 888 |
| 93 | 656 | 7 | 756 |
| 92 | 733 | 2 | 827 |
| 77 | 499 | 3 | 579 |
| 144 | 541 | 11 | 696 |
| 111 | 702 | 2 | 815 |
| 188 | 1130 | 0 | 1318 |
| 165 | 1075 | 0 | 1239 |
| 76 | 997 | 1 | 1074 |
| 62 | 648 | 2 | 711 |
| 65 | 675 | 1 | 741 |
| 63 | 894 | 0 | 957 |
| 46 | 805 | 15 | 866 |
| 55 | 764 | 15 | 833 |

Table 7.8.2. Nephrops in FU19 (SW and SE Ireland). Irish Nephrops directed effort hrs and lpue, 1993-2009.

| Year | Irish Fleet |  |  |
| :---: | :---: | :---: | :---: |
|  | Nephrops trawlers (>30\% landings weight) |  |  |
|  | Effort hrs | Landings Tonnes | LPUE Kg/hr |
| 1995 | 9126 | 206 | 22.5 |
| 1996 | 9295 | 220 | 23.7 |
| 1997 | 9604 | 248 | 25.8 |
| 1998 | 15775 | 386 | 24.5 |
| 1999 | 13345 | 206 | 15.4 |
| 2000 | 9329 | 178 | 19.1 |
| 2001 | 9701 | 309 | 31.8 |
| 2002 | 25565 | 764 | 29.9 |
| 2003 | 28887 | 621 | 21.5 |
| 2004 | 26554 | 529 | 19.9 |
| 2005 | 23848 | 455 | 19.1 |
| 2006 | 24272 | 460 | 19.0 |
| 2007 | 30361 | 665 | 21.9 |
| 2008 | 25101 | 573 | 22.8 |
| 2009 | 22797 | 527 | 23.1 |

Table 7.8.3. Nephrops in FU19 (SW and SE Ireland). Mean time-series for catches and landings.

| Year | Catches |  | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | Females | Males | Females | Males | Females |
|  | Ma | na | na | na | na | na |
| 1995 | na | 31.1 | 29.7 | 38.7 | 38.8 |  |
| 1996 | 34.5 | 31.3 | 31.2 | 30.9 | 39.8 | 38.4 |
| 1997 | 34.6 | 32.9 | 31.2 | na | na | na |
| 1998 | na | na | na | na | 31.2 | 41.3 |
| 1999 | 38.5 | 35.4 | 31.8 | 39.1 |  |  |
| 2000 | na | na | na | na | na | na |
| 2001 | na | na | na | na | na | na |
| 2002 | 30.4 | 28.8 | 29.7 | 28.8 | 39.9 | 40.5 |
| 2003 | 33.1 | 29.4 | 31.1 | 30.0 | 38.4 | 38.0 |
| 2004 | 32.8 | 28.8 | 32.0 | 30.2 | 39.8 | 37.7 |
| 2005 | 31.3 | 27.5 | 29.1 | 26.9 | 38.4 | 37.0 |
| 2006 | 34.4 | 31.7 | 31.4 | 30.4 | 38.9 | 37.7 |
| 2007 | 35.6 | 33.2 | 32.4 | 31.7 | 39.1 | 38.2 |
| 2008 | 36.2 | 33.1 | 32.5 | 31.6 | 38.9 | 38.1 |
| 2009 | 33.9 | 29.2 | 31.2 | 29.8 | 39.3 | 37.4 |
| na not available |  |  |  |  |  |  |

Table 7.8.4 Nephrops in VII summary table of landings by Function Unit and outside FU for TAC Area VII.

|  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |



Figure 7.8.1. Nephrops in FU19 (Ireland SW and SE Coast). The spatial distribution of the fishery of the Irish Fishery from VMS data.


Figure 7.8.2. Nephrops in FU19 (Ireland SW and SE Coast). Landings in tonnes by country.


Figure 7.8.3. Nephrops in FU19 (Ireland SW and SE Coast). Trawl effort for Irish OTB vessels where $\mathbf{> 3 0 \%}$ of landed weight was Nephrops. Trawl lpue for Irish OTB vessels where $\mathbf{> 3 0 \%}$ of landed weight was Nephrops.


Figure 7.8.4. Nephrops in FU19 (Ireland SW and SE Coast). Mean size trends for catches and whole landings by sex.

### 7.9 Plaice in West of Ireland Division VII b, c

Type of assessment in 2010
No assessment was performed.

### 7.9.1 General

## Stock Identity

Plaice in VIIb are mainly caught by Irish vessels on sandy grounds in coastal areas. Plaice catches in VIIc are negligible. There are two distinct areas in which plaice are caught by Irish vessels in VIIb: an area to the west of the Aran Islands and an area in the north of VIIb which extends into VIa (the Stags Ground). During 1995-2000 a large proportion of the VIIbc plaice landings were taken from the Stags Grounds (Rectangles 37D8, 37D9, 37E0 and 37E1). The landings and lpue in this area have dropped sharply since 2000, in line with a general decrease of lpue in Division VIa. The landings and lpue on the Aran grounds appear to have been more or less stable since the start of the logbooks time-series in 1995 (WD 1, WGCSE 2009). It is not known how much exchange there is between plaice on the Aran grounds and those on the Stags ground.

### 7.9.2 Data

The nominal landings are given in Table 7.9.1.

Table 7.9.1. Landings of plaice in VIIbc as officially reported to ICES.

| Country | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | 2 | - | - | - | - | - | - |
| France | 60 | 45 | 10 | 9 | 4 | 16 | 6 | 12 | 9 | 8.00 | 37 | 2 | 10 |
| Ireland | 124 | 106 | 153 | 133 | 135 | 122 | 117 | 142 | 135 | 122 | 108 | 110 | 150 |
| Spain | - | - | - | - | - | - | - | 65 | 58 | 22 | 7 | - | - |
| UK - Eng+Wales+N.IIr. | . | . | . | . | . | . |  |  |  |  |  |  |  |
| UK - England \& Wales | 1 | 1 | - | - | - | - | - | - | 4 | 4 | - | 3 | 7 |
| UK - Scotland | - | - | - | - | - | - | - | - | - | - | - | 3 | - |
| Total | 185 | 152 | 163 | 142 | 139 | 138 | 125 | 219 | 206 | 156 | 152 | 118 | 167 |


| Country | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | - | - | - | - | - | - | - |
| France | 11 | 13 | 9 | 1 | 11 | 9 | 3 | 2 | 1 | 5 | 1 | 3 | - |
| Ireland | 114 | 153 | 157 | 159 | 130 | 179 | 180 | 191 | 200 | 239 | 248 | 206 | 160 |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - |
| UK - Eng+Wales+N.IIr. |  |  |  | 1 | 2 | - | 6 | 1 | 2 | 1 | 2 | - | 1 |
| UK - England \& Wales | 5 | 1 | 2 |  |  |  |  |  |  |  | . | . | . |
| UK - Scotland | - | - | - | 13 | 90 | 3 | 3 | 2 | 3 | 1 | - | - | - |
| Total | 130 | 167 | 168 | 174 | 233 | 191 | 192 | 196 | 206 | 246 | 251 | 209 | 161 |
| Unallocated |  |  |  |  |  |  |  |  |  |  | -11 | 4 | 22 |
| WG estimate |  |  |  |  |  |  |  |  |  |  | 240 | 213 | 183 |
| Country | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| Denmark | - | - | - | - |  |  |  |  |  |  |  |  |  |
| France |  | 31 | 8 | 17 | 7 | 14 | 12 | 11 | 12 | 9 |  |  |  |
| Ireland | 157 | 99 | 70 | 51 | 56 | 39 | 25 | 20 | 23 | 21 | 20 |  |  |
| Spain | - | - | - | 2 |  |  |  | 1 |  | 1 |  |  |  |
| UK - Eng+Wales+N.IIr. | - | - | - | 2 |  | 0 | 0 | 0 |  |  |  |  |  |
| UK - England \& Wales |  | . | . | . |  |  |  |  |  |  |  |  |  |
| UK - Scotland | 2 | - | - | - | 0 |  |  |  |  |  |  |  |  |
| UK |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 159 | 130 | 78 | 72 | 63 | 53 | 37 | 31.6 | 35.3 | 31 | 20 |  |  |
| Unallocated | 13 | -22 | 9 | -1 | 9 | 2 | 1 | -1 | -1 | 4 |  |  |  |
| WG estimate | 172 | 108 | 87 | 71 | 72 | 55 | 38 | 30 | 34 | 35 |  |  |  |

### 7.10 Plaice in Divisions VIIf,g (Celtic Sea)

## Type of assessment in 2010

Update, no changes to the assessment.

## ICES advice applicable to 2010

ICES advises on the basis of exploitation boundaries in relation to precautionary considerations that a $50 \%$ reduction in $F$ is needed to increase SSB to around $B_{p a}$ in 2011. This corresponds to landings of less than 330 tin 2010.

## ICES advice applicable to 2009

Exploitation boundaries in relation to precautionary considerations: a $75 \%$ reduction in $F$ is needed to increase SSB to around Bpa in 2010. This corresponds to landings of less than $170 t$ in 2009.

### 7.10.1 General

## Stock description and management units

A TAC is in place for ICES Areas VIIf\&g which corresponds to the stock area.

## Management applicable to 2009 and 2010

TACs and quotas set for 2009

| Species: Plaice <br> Plaurnenetes platess |  | Zone: | VIIf and VIIg (PLE] 7 FG.) |
| :---: | :---: | :---: | :---: |
| Belgium | 59 |  |  |
| France | 107 |  |  |
| Ireland | 200 |  |  |
| United Kingdom | 56 |  |  |
| EC | 422 |  |  |
| TAC | 422 |  | Analytical TAC <br> Article 3 of Regulation (EQ) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TACs and quotas set for 2010

| Species: | Plaice <br> Pleuronectes platessa | Zone:VIIf and VIIg <br> (PLE/7FG) |
| :--- | :--- | :--- | :--- |
| Belgium | 67 |  |
| France | 120 |  |
| Ireland | 201 |  |
| United Kingdom | 63 | Analytical TAC |
| EU | 451 |  |
| TAC | 451 |  |

## Fishery in 2009

The main fishery is concentrated on the Trevose Head ground off the north Cornwall coast and around Land's End. Although plaice are taken throughout the year, heavi-
est landings are in March, after the peak of spawning, with a second peak in September. The fisheries taking plaice in the Celtic Sea mainly involve vessels from Belgium, France, England and Wales. In 2008 Belgium reported $50 \%$ of officially reported landings, the UK $17 \%$, France $17 \%$, and Ireland the remaining $17 \%$. In 2009 , officially reported French data were not available; of the landings reported Belgium reported $58 \%$, the UK $15 \%$ and Ireland $17 \%$. The WG estimated total international landings for 2009 were $463 \mathrm{t}, 10 \%$ above the TAC ( 420 t ) and $21 \%$ below the status quo prediction given by last year's assessment ( 580 t ).

Although the current assessment indicates a decline in F in recent years, it is unclear as to whether this is linked to the Trevose Head spring fishery closure. Discards are considered to be significant but are presently not quantified.

### 7.10.2 Data

## Landings

National landings data and estimates of total landings used by the WG are given in Table 7.10.1. Minor revisions were reported to landings data for 2008.

## Discards

Indications are that discard rates, although variable, may be substantial in some fleets/periods. Total raised discard information is available for some fleets, and data raised to sampled vessels for others, but the WG has not yet been able to analyse these data. For this update assessment, discard data were excluded pending a more thorough examination at the next Benchmark Workshop. All references to 'catch' (e.g. 'catch weights-at-age, etc.) therefore relate to landings only.

Data from 2009 discard sampling programmes are summarized in Figures 7.10.3a and b.

## Biological

Annual length compositions for 2009 are given in Table 7.10.4, and length compositions for $\mathrm{UK}(\mathrm{E}+\mathrm{W})$ landings for the last ten years are presented in Figure 7.10.4.

Following minor revisions to landings data for previous years (see above), the international age compositions and weights-at-age have been amended.

Quarterly age compositions for 2009 were available for Belgium, Ireland and UK (E+W), representing approximately $72 \%$ of the total landings. Methods for the derivation of international catch numbers-at-age and for the calculation of catch and stock weights-at-age are fully described in the Stock Annex, Section B. 1 and B.2.

Parameter estimates for the in-year smoothing of catch and stock weights in 2009 are as follows:

$$
\mathrm{Wt}=0.012 \mathrm{Age}^{2}-0.092 \mathrm{Age}+0.447 \quad\left(\mathrm{R}^{2}=0.91\right)
$$

Catch weights-at-age are plotted as mid-year, and stock weights-at-age are interpolated from the fitted curve at 1 January. The catch and stock weights-at-ages 1 and 2 for 2008 and 2009 exhibited bias due to the application of the quadratic smoother, they were modelled as being heavier than ages 3 and 4 . In order to correct the bias average weights-at-age over the years 2002-2007 were used for ages 1 and 2 in 2008 and 2009.

The age compositions of landings for the last 10 years are shown in Figure 7.10.5. Catch numbers and weights-at-age in the catch and stock as used for the assessment are given in Tables 7.10.5-7.10.7. As in previous assessments, numbers-at-age 1 have been replaced by zero values; see Section B. 1 in the Stock Annex.

A natural mortality estimate of 0.12 was applied to all ages and to all years, as previously. The maturity ogive used in this update assessment was derived in 1997 and was applied to all years. Further details of the derivation of these can be found in Section B. 2 of the Stock Annex.

## Surveys

Indices of abundance from the UK (BTS-Q3) beam trawl survey in VIIf and the Irish Celtic Explorer IBTS survey (IBTS-EA-4Q) are presented in Table 7.10.8. The UK (E\&W) data indicate relatively strong 1994 and 1999 year classes. The Celtic Explorer IBTS survey-series started in 2003 and is not yet included in the assessment.

Figure 7.10.6 presents the log UK (BTS-Q3) cpueUE indices by year and year class, the $\log$ catch curves for each cohort and the gradient of the catch curves used as an indication of total mortality trends. The plots illustrate the historical consistency of year class estimates from the survey, with less agreement in recent years.

## Commercial cpue

Commercial tuning indices of abundance from the UK (E\&W) beam trawl and otter trawl data are presented in Table 7.10.9. Figures 7.10.7a and b presents the log commercial cpue indices by year and year class, the log catch curves for each cohort and the gradient of the catch curves used as an indication of total mortality trends. The plots illustrates the historical consistency of year class estimates from the commercial data throughout the time-series for the beam trawls with more noise resulting from two major year effects in the otter trawl data.

Effort and lpue data were available for the UK (E+W) beam trawl, UK (E\&W) otter trawl, Irish otter trawl, beam trawl and seine fleets, Belgian beam trawl and the UK September beam trawl survey (Tables 7.10.2, 7.10.3 and Figures 7.10.1, 7.10.2).

Commercial lpue data appear to show a general pattern of steep decline since the high levels in the early 1990s, with a further decline in recent years. There was an increase in 2007 and 2008 for beam trawlers in VIIf and a smaller increase in 2007 and 2008 for otter trawlers in VIIg east but the levels returned to the recent low levels in 2009.

UK (E\&W) beam trawl effort levels have declined in both VIIf and VIIg from the high levels observed in 1999-2001; effort in VIIf in 2009 was at the lowest level since 1983. UK (E\&W) otter trawl effort levels for VIIf and VIIg have shown a general decline since 1990, increased in VIIf after 2000 and have been relatively stable since 2003.

Irish otter trawl effort has steadily increased since 1999, while beam trawl show a less pronounced increase over the time-series prior to 2008, with a decrease in 2008 and 2009; the Irish seine fleet shows only a weak downward since 2003.

## Other relevant data

Other than the rectangle closures, there were no early closures of the fishery for plaice in 2009. There is relatively little information on the level of landings misreporting on this stock, although it is not considered to be a problem. Reports from industry suggest that the main issues affecting the fishery in VIIf\&g are displacement of effort
due to the rectangle closures and the restrictions on the use of 80 mm mesh west of $7^{\circ} \mathrm{W}$.

### 7.10.3 Stock assessment

Section 1.4.1 outlines the general approach adopted at this year's Working Group meeting, and the specific approach for this stock is given in the Stock Annex.

## Data screening

A separable analysis was carried out to screen the catch-at-age data; no anomalies were apparent (results are available in the 'Exploratory runs' folder).

## Final update assessment

The commercial tuning data available, and the subset used in the assessment, are given in Table 7.10.9. No exploratory XSA runs were carried out for this update assessment.

Final settings, used since 2005, are detailed below:

| 2009 XSA |  |  |
| :---: | :---: | :---: |
| Fleets UK-CSBT | 90-09 | 4-8 |
| UK-CSOT | 89--09 | 4-8 |
| UK-BCCSBTS-S | 90-09 | 1-5 |
| Taper |  |  |
| Taper range |  | - |
| Ages catch dep. Stock size |  | 1-5 |
| Q plateau |  | 7 |
| F shrinkage se |  | 2.5 |
| year range |  | 5 |
| age range |  | 4 |
| Fleet threshold se |  | 0.5 |
| Age range |  | 1-9+ |
| Age 1 catch numbers |  |  |
| Fbar age range |  | 3-6 |

XSA diagnostics from the final run are given in Table 7.10.10 and log catchability residuals plotted in Figure 7.10.8. Survivor estimates for ages 4 and above are reasonably consistent between fleets. The standard error threshold operates on the commercial fleet data, maintaining relatively even weighting up to age 5 , after which the survey contribution sharply declines reflecting the lack of data for older ages. The survey contributes around $40 \%$ of the weight to estimates of survivors at age 1, with the remainder coming from P-shrinkage. At ages 2 and 3 the survey provides around $65 \%$ of the weighting. F-shrinkage is negligible throughout.

The residuals for the UK beam trawl survey (Figure 7.10.8) show apparent cohort and age effects, in contrast to the commercial data sets for older ages. The Working Group considers this to be due to the heavy discarding of the youngest age classes, resulting in these fish being absent from commercial data. The survey estimates of these age classes are probably a true reflection of their strength within the fishery. There are some year effects apparent in the otter trawl fleet data, but these are relatively small.

The retrospective analysis (Figure 7.10.9) shows a historical tendency to underestimate terminal F values, and to overestimate SSB, recent estimates are based in the same direction but more consistent. There is a pattern of overestimating recruitment, which is likely to result from discarding causing overestimation of the future landings numbers at the youngest ages and partly due to the use of population shrinkage, which will tend to bias recruit estimates towards the mean during periods of low recruitment. Recent estimates of recruitment have been more stable from year to year.

Fishing mortalities and population numbers from the final XSA run are given in Tables 7.10.11 and 7.10.12 and the summary in Table 7.10.13. Fishing mortality in 2009 is estimated to have been at 0.41, which is the lowest in the time-series. SSB in 2009 is estimated at 1128 t above $\mathrm{B}_{\lim }(1100 \mathrm{t})$ for the first time since 2001 . However, the retrospective analysis suggests that this estimate of SSB is likely to be an overestimation.

## Comparison with previous assessments

The current assessment suffers from a retrospective pattern, when compared with results from previous assessments, which is evident in SSB, fishing mortality and recruitment (Figure 7.10.13).

## State of the stock

A summary of the time-series of XSA results is given in Table 7.10.13 and Figure 7.10.10. Fishing mortality has fluctuated without trend since 1977 but has declined since 2004. The most recent value should be viewed with caution as the retrospective pattern implies that these may be revised upwards in next year's assessment.

SSB rose to a high level throughout the 1980s, as a result of a series of above-average recruitments, but has declined since 1990. SSB is estimated to have been below $\mathrm{B}_{\mathrm{pa}}$ ( 1800 t) since 1996, and SSB was below $B_{\lim }(1100$ t) between 2002 and 2008. In 2009, SSB is estimated to be 1128 t , at Blim, but this initial estimate is typically revised downward by $\sim 10 \%$ in the following year's assessment due to the retrospective pattern present in the stock (Figure 7.10.9).

With the exception of the 1994 year-class, all recruitments-at-age 1 since 1992 have been below the long-term arithmetic average ( 4.4 million).

### 7.10.4 Short-term projections

## Estimating year-class abundance

The XSA estimate of the 2007 year class ( 2.2 million 1 year olds in 2008) has been revised downwards from last year's estimate ( 2.9 million). The 2008 year class is estimated at 3.0 million 1 year olds in 2009; the sources of this estimate are detailed in the table below. Recruitment estimates for subsequent years were derived from a shortterm geometric mean ( $\mathrm{GM}_{1989-08}, 2.8$ million).

Working Group estimates of year-class strength used for prediction can be summarised as follows:

Recruitment-at-age 1:

| Year class | Thousands | Basis | Surveys | Commercial | P Shrinkage |
| ---: | :---: | :--- | :---: | :---: | :---: |
| 2007 | 2191 | XSA | $65 \%$ | - | $35 \%$ |
| 2008 | 3021 | XSA | $44 \%$ |  | $56 \%$ |
| 2009 | 2815 | GM $(89-08)$ |  |  |  |
| 2010 | 2815 | GM (89-08) |  |  |  |

The input values for the catch forecast (using the MFDP software) are given in Table 7.10.14. The F at age values used were calculated as the mean of the XSA values from 2007-2009, unscaled. Catch and stock weights-at-age were also the mean of the period 2007-2009. Stock numbers-at-age in 2009 for ages 2 and older were obtained from the XSA. SSB values are calculated for 1 January.

Table 7.10.15 gives the management option table from the status quo catch prediction, and short-term results are shown in Figure 7.10.12. Assuming status quo $\mathrm{F}\left(\mathrm{F}_{\text {sq }}=0.44\right)$ implies landings of 539 t in 2010 and 556 t in 2011. (The TAC for 2010 is 451 t .). SSB is predicted to remain stable at 1300 t in 2010 and 1350 t in 2011, increasing to 1400 t in 2012. These results are discussed further in Section 7.10.10.

The detailed output for the status quo F forecast by age group is given in Table 7.10.16, and the estimated contributions of recent year classes to the predicted catches and SSBs are given in Table 7.10.17. The assumptions of GM1989-08 recruitment are predicted to contribute 7\% to the landings in 2011 and 28\% to SSB in 2012.

The stock and recruitment scatterplot is given in Figure 7.10.11; it should be recalled that the unknown mortality from discarding would have an impact on the stockrecruitment relationship. It is noteworthy that recruitment is strongly auto-correlated for this stock such that there is a dominant recruitment - SSB relationship rather than stock and recruitment; recruitment has exhibited strong and weak series at high stock biomass.

The recruitment of Celtic Sea plaice and neighbouring stocks appear to respond to negatively to sea temperature anomalies (Fox et al., 2000). However, compared with fisheries on some other species (e.g. cod), those on plaice are less dependent on the incoming year class. Therefore, the incorporation of temperature data into plaice stock dynamic models will probably not have a large impact on short-term projections, but may allow medium to long-term forecasts to be made under varying environmental scenarios.

### 7.10.5 Maximum sustainable yield evaluation

The VIIf\&g plaice times-series of assessment stock and recruit estimates, fishing mor-tality-at-age (average of the most recent three year), catch and stock weights (10 year averages), maturity and natural mortality-at-age were used to estimate proxies for the fishing mortality biomass and landings at maximum sustainable yield ( $\mathrm{F}_{\mathrm{msy}} \mathrm{B}_{\mathrm{B}} \mathrm{B}_{\text {my }}$ and MSY) within the srmsymc program. The sen and sum input data files are presented in Tables 7.10.19 and 7.10 .20.

Three stock and recruit models are fitted by the program, Ricker, Beverton and Holt and the smooth hockey stick Figures 7.10.13-7.10.16. Based on the A.I.C. all models have an equal fit to the available data. The estimates of $\mathrm{F}_{\text {crash, }} \mathrm{F}_{\text {msy }}, \mathrm{B}_{\text {msy }}$ and MSY are presented with their percentiles and coefficients of variation in Table 7.10.21. Figure 7.10.17 illustrates the uncertainty in yield-per-recruit curve, with estimates also presented in Table 7.10.21.

Each model assumes that there is a relationship between increasing levels of recruitment and increasing spawning-stock. However, as shown in Figure 7.10.11 and, as discussed in the previous section, for VIIfg plaice recruitment has exhibited strong autocorrelation and the dominant determinant of recruitment is likely to be environmental conditions rather than the level of SSB.

The link to environmental control of recruitment and independence from SSB would suggest the use of yield-per-recruit fishing mortality reference levels as appropriate. However, as shown by Figure 7.10 .17 the form of the YPR curve is poorly determined and the estimates of $\mathrm{F}_{0.1}, \mathrm{~F}_{35 / 40 \%} \mathrm{SPR}$ have high cv, with $\mathrm{F}_{\text {max }}$ poorly determined (Table 7.10.21).

For all stock and recruit relationships, yield is forgone at levels of fishing mortality below the estimates of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{35 \%}$ and as F increases yield is maximised at 0.19 for the Beverton and Holt model, 0.36 for the Ricker model and 0.34 for the smooth hockey stick. All of the estimates of $\mathrm{F}_{\text {msy }}$ are below the 5 th percentiles of $\mathrm{F}_{\text {crash }}$ ( 0.53 , $0.57,0.43$ respectively).

Estimates of $\mathrm{F}_{\text {msy }}$ differ between the assumptions for the stock and recruitment dynamics of plaice in VIIfg, each are equally plausible and there is no way of distinguishing between them. Consequently WGCSE consider that fishing mortalities in the range $0.19-0.36$ are consistent with maximising long-term yield for plaice in VIIfg.

## Yield-per-recruit analysis

Results for deterministic yield and SSB per recruit (using program MFYPR), conditional on the recent exploitation pattern, are given in Table 7.10.18 and Figure 7.10.12. $\mathrm{F}_{\max }$ is given by a reference F of 0.35 , around $80 \%$ of $\mathrm{Fsq}_{\text {s }}$. Long-term yield and SSB (at $\mathrm{F}_{\mathrm{sq}}$ and assuming $\mathrm{GM}_{89-08}$ recruitment $=2.8$ million) are given as 610 t and 1538 t respectively.

### 7.10.6 Precautionary approach reference points

The Working Group's current approach to reference points is outlined in Section 1.4.4. Current reference points are detailed below.

| $\mathrm{F}_{\lim }$ | No proposal |  |  |
| :--- | ---: | :--- | :--- |
| $\mathrm{F}_{\mathrm{pa}}$ | No proposal |  |  |
| $\mathrm{B}_{\lim }$ | 1100 t | Basis | $\mathrm{B}_{\text {loss }}(\mathrm{B} 78, \mathrm{WG} 98)$ |
| $\mathrm{B}_{\mathrm{pa}}$ | 1800 t | basis | $\mathrm{Blim}_{\text {lim }} \cdot \exp \left(1.645^{*} 0.3\right)$ |

SSB is currently below $\mathrm{B}_{\mathrm{pa}}$ and at $\mathrm{F}_{\mathrm{sq}}$ will remain below $\mathrm{B}_{\mathrm{pa}}$ in the short term.
Bloss (1010 t, 1978) is considered stable. The estimate of SSB in 1978 has not been revised in the last nine assessments of this stock. Further details can be found in Section G of the Stock Annex. A general discussion on target reference points is given in Section 1.

### 7.10.7 Management plans

There is no management plan for Celtic Sea plaice.
This WG has in the past provided a number of scenarios for potential management plans for Celtic Sea plaice (ICES, 2006). The analyses indicated that an F in the range 0.25 to 0.56 would be sustainable in terms of maintaining the stock above $B_{p a}$. The
range was also considered consistent with sustainable fishing for VIIfg sole. The range of fishing mortalities estimated within the MSY framework overlap with the lower portion of the previous management plan range due to the requirement to maximise yield

### 7.10.8 Uncertainties in assessment and forecast

## Sampling

Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches, and associated CVs of some national catch-at-age datasets are available in the Stock Annex. The sampling levels for those countries supplying information are given in Table 2.1.

## Discards

Estimates of discarding are not included in this assessment. However, data from discard sampling indicates that rates are high for this stock in some seasons/fleets (Figures 7.10.3a, b) and their non-inclusion may represent a major deficiency in the assessment. The composition of the fleets and therefore the gear types employed in the fishery show fluctuations over time, so it is likely that the discard rates observed in the fishery now are not applicable to periods earlier in the time-series. From 2003 onwards, discard sampling for Ireland, Belgium, France and UK (E\&W) has been improved under the Data Collection Regulation.

## Consistency

The trends and estimates of fishing mortality, SSB and recruitment in this assessment are consistent with last year's assessment (Figure 7.10.13). Last year's assessment estimated F in 2008 at 0.365 ; this year it has been revised upwards to 0.41 . Last year's assessment estimated SSB in 2008 to be 1243 t ; this estimate has been revised downwards to around 1063 t this year.

## Misreporting

Misreporting has been considered a potential problem for this stock in earlier years. However, misreporting of catches across ICES Divisions is thought to be minor. The status quo forecast indicates landings around $20 \%$ in excess of the TAC for 2010 . It should be noted that even though total reported landings for this stock in recent years have been below the TAC, fleets may be restricted by their individual quota allocation.

### 7.10.9 Recommendation for next Benchmark

| Year | Candidate Stock | Supporting Justification | Suggested time | Indicate expertise necessary at benchmark meeting. |
| :---: | :---: | :---: | :---: | :---: |
| 2009 | VIIf,g <br> Plaice | Biological interactions with adjacent stocks; in particular Irish Sea plaice (Dunn and Pawson, 2002) | 2011 | Expert Group members |
|  |  | The lack of discard data in the assessment |  |  |
|  |  | The need to review the tuning data included in the assessment as additional survey data are becoming available and the age ranges of all tuning datasets should be reinspected |  |  |
|  |  | Need to reconsider the assessment model setting being used as the power model is only indicated by one of the fleets |  |  |
|  |  | There is also a retrospective bias in SSB, F and recruitment, which should be examined |  |  |
|  |  | Review the new maturity data, which has become available from sampling carried out under the EU DCR |  |  |

### 7.10.10 Management considerations

The SSB of this stock is estimated to have been below $\operatorname{Blim}(1100 t)$ since 2002, but the SSB is now estimated to have been 1128 t in 2009. However, the retrospective plot indicates that this is expected to be revised downwards next year by $\sim 10 \%$. The status quo catch forecast implies that SSB will continue rise above Blim to 1300 t in 2010, 1350 t in 2011 and 1400 t in 2012, assuming GM89-08 recruitment levels. Despite the use of a recent GM recruitment level, catch forecasts in recent years have been overlyoptimistic, e.g. last year's forecast indicated landings of 580 t in 2009, whereas the landings value used by the WG this year was 463 t . The level of fishing mortality in the last three years appears to have been at the lowest level for the time-series, which is consistent with the reductions in effort in the beam trawl fleets. Landings in 2010 are this year predicted to be 540 t , well above the TAC of 450 t .

The high level of discarding indicated for some fleets in this fishery would suggest a mismatch between the mesh size employed in the fishery and the size of the fish being landed on the market. Increases in the mesh size of the gear should result in fewer discards and ultimately, in increased yield from the fishery. The results of studies presented to the 2004 WG (ICES, 2004) indicate that this would also benefit the sole VIIf,g stock without decreasing sole landings in the long term.

## Regulations and their effects

Technical measures in force for this stock are minimum mesh sizes, minimum landing size, and restricted areas for certain classes of vessels. Technical regulations regarding allowable mesh sizes for specific target species, and associated minimum landing sizes, came into force on 1 January 2000 (Section 2.1). The minimum landing size for plaice in Divisions VIIf, g is currently 27 cm .

Since 2005, ICES rectangles 30E4, 31E4, and 32E3 have been closed during the first quarter (Council Regulations 27/2005, 51/2006, 41/2007, 40/2008, 43/2009 and 2010) with the intention of reducing fishing mortality on cod. There is evidence that this closure has redistributed effort to other areas. Many vessels (particularly beam trawlers from the UK and Belgium) fished close to the borders of the closed rectangles during the closure, and fished intensively inside the rectangles when they were reopened. Information from the UK shows that plaice can be caught in areas outside of the closed area with the same catch rates. Fishing mortality has decreased since 2005, and the closure may have been one of the contributing factors.

## References

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ICES, 2004. Report of the Working Group on the Assessment of Southern Shelf Demersal Stocks (WGSSDS), 29 June-8 July 2004, Oostende, Belgium. ICES CM 2005/ACFM:03. 543 pp.

Table 7.10.1. Plaice in Divisions VIIf\&g. Nominal landings (t) as reported to ICES, and total landings as used by the Working Group.
National landings as estimated by the working group 1977-1985; as reported to ICES and total landings as used by the working group 1986 onwards

|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 214 | 196 | 171 | 372 | 365 | 341 | 314 | 283 | 357 | 665 | 581 | 617 | 843 | 794 | 836 | 371 |
| UK (Engl. \& Wales) | 150 | 152 | 176 | 227 | 251 | 196 | 279 | 366 | 466 | 529 | 496 | 629 | 471 | 497 | 392 | 302 |
| France | 365 | 527 | 467 | 706 | 697 | 568 | 532 | 558 | 493 | 878 | 708 | 721 | 1089 | 767 | 444 | 504 |
| Ireland | 28 | 0 | 49 | 61 | 64 | 198 | 48 | 72 | 91 | 302 | 127 | 226 | 180 | 160 | 155 | 180 |
| N . Ireland |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  |  |  |
| Netherlands |  |  |  |  |  |  |  |  |  | 9 |  |  |  |  |  |  |
| Scotland | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 1 |  |  |  | 1 |  | 5 |
| Total | 757 | 875 | 863 | 1373 | 1377 | 1303 | 1173 | 1279 | 1407 | 2384 | 1912 | 2194 | 2583 | 2219 | 1827 | 1362 |
| Unallocated | 0 | 0 | 0 | 0 | 0 | 0 | -27 | -69 | 345 | -693 | -11 | -78 | -432 | -137 | -326 | -174 |
| Total as used by WG | 757 | 875 | 863 | 1373 | 1377 | 1303 | 1146 | 1210 | 1752 | 1691 | 1901 | 2116 | 2151 | 2082 | 1501 | 1188 |



Table 7.10.2. Plaice in Divisions VIIf\&g. Lpue for UK (E\&W) fleets.

|  | LANDINGS PER UNIT EFFORT (LPUE) |  |  |  |  |  | LANDINGS/EFFORT DATA <br> RECT GROUP VIIf (Grp1) |  |  |  | ADDITIONAL EFFORT DATA |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RECT. GROUP VIIf (grp 1) |  | RECT. GROUP VIIg EAST (grp 2) |  | RECT. GROUP VIIg WEST (grp 3) |  |  |  |  |  | VIlg (East) |  | VIIg (West) |  |
|  |  |  | otter trawl catch | VIIf (grp1) Beam trawl catch |  | Otter | Beam | Otter | Beam |
|  | TRAWL | $\begin{gathered} \hline \text { BEAM } \\ \text { TRAWL } \\ \hline \end{gathered}$ |  |  | TRAWL | $\begin{gathered} \hline \text { BEAM } \\ \text { TRAWL } \\ \hline \end{gathered}$ | TRAWL | $\begin{gathered} \hline \text { BEAM } \\ \text { TRAWL } \\ \hline \end{gathered}$ | tonnes | $\begin{gathered} 000 \mathrm{~s} \\ \text { hr fished } \\ \hline \end{gathered}$ | tonnes | $\begin{gathered} 000 \mathrm{~s} \\ \text { hr fished } \end{gathered}$ | $\begin{gathered} 000 \mathrm{~s} \\ \text { hr fished } \end{gathered}$ | $\begin{gathered} 000 \mathrm{~s} \\ \mathrm{hr} \text { fished } \\ \hline \end{gathered}$ | $\begin{gathered} 000 \mathrm{~s} \\ \text { hr fished } \\ \hline \end{gathered}$ | $\begin{array}{\|c\|} 000 \mathrm{~s} \\ \mathrm{hr} \text { fished } \end{array}$ |
| 1972 | 7.70 |  | 4.97 |  |  |  | 1.15 |  | 361.82 | 45.72 |  |  | 6.01 |  | 0.74 |  |
| 1973 | 7.54 |  | 2.75 |  | 34.92 |  | 353.95 | 45.28 |  |  | 3.59 |  | 0.05 |  |
| 1974 | 4.99 |  | 1.22 |  | 0.00 |  | 198.12 | 38.94 |  |  | 2.03 |  | 0.00 |  |
| 1975 | 4.88 |  | 4.07 |  | 0.75 |  | 173.01 | 33.53 |  |  | 10.35 |  | 0.04 |  |
| 1976 | 4.54 |  | 2.70 |  | 2.13 |  | 112.09 | 25.61 |  |  | 5.21 |  | 0.04 |  |
| 1977 | 4.06 |  | 1.76 |  | 0.00 |  | 102.81 | 27.16 |  |  | 5.36 |  | 0.04 |  |
| 1978 | 4.19 | 3.06 | 2.24 | 0.00 | 0.00 | 0.00 | 117.74 | 27.08 | 7.58 | 2.50 | 6.73 | 0.00 | 0.00 | 0.00 |
| 1979 | 5.31 | 3.62 | 3.34 | 2.19 | 0.00 | 0.00 | 125.81 | 23.84 | 6.30 | 1.96 | 4.54 | 0.13 | 0.00 | 0.00 |
| 1980 | 5.91 | 4.27 | 4.03 | 7.15 | 2.46 | 0.00 | 162.29 | 26.43 | 17.65 | 4.31 | 2.67 | 0.10 | 0.60 | 0.00 |
| 1981 | 5.36 | 3.50 | 3.20 | 3.13 | 1.05 | 5.23 | 126.27 | 24.10 | 23.72 | 6.24 | 7.78 | 0.78 | 4.78 | 0.10 |
| 1982 | 4.82 | 5.10 | 1.14 | 6.73 | 0.06 | 5.57 | 92.65 | 19.20 | 55.42 | 9.95 | 7.50 | 1.86 | 2.56 | 0.58 |
| 1983 | 6.05 | 3.92 | 2.66 | 5.24 | 0.00 | 4.88 | 108.76 | 17.61 | 47.72 | 12.35 | 5.33 | 6.82 | 0.00 | 0.80 |
| 1984 | 6.15 | 6.41 | 4.90 | 7.49 | 0.00 | 4.14 | 160.64 | 23.16 | 99.01 | 13.55 | 4.35 | 4.31 | 0.00 | 2.06 |
| 1985 | 6.98 | 6.38 | 5.09 | 8.05 | 2.61 | 7.10 | 188.06 | 25.24 | 146.73 | 18.69 | 5.72 | 5.14 | 0.57 | 1.41 |
| 1986 | 6.62 | 5.22 | 4.28 | 10.62 | 1.44 | 11.31 | 142.84 | 21.18 | 90.44 | 20.72 | 7.72 | 4.31 | 0.82 | 0.68 |
| 1987 | 6.60 | 4.32 | 6.46 | 10.79 | 0.86 | 10.66 | 199.03 | 24.43 | 145.37 | 38.76 | 9.87 | 4.83 | 0.83 | 0.92 |
| 1988 | 10.04 | 8.53 | 7.32 | 9.95 | 1.97 | 14.42 | 205.56 | 20.09 | 204.58 | 25.62 | 9.96 | 2.18 | 0.43 | 0.88 |
| 1989 | 7.40 | 5.63 | 6.36 | 9.67 | 4.35 | 16.42 | 130.67 | 17.61 | 96.05 | 20.26 | 8.13 | 3.72 | 0.25 | 0.26 |
| 1990 | 4.16 | 3.93 | 2.43 | 6.80 | 2.70 | 5.34 | 97.82 | 22.56 | 157.15 | 30.77 | 10.55 | 4.89 | 0.45 | 4.32 |
| 1991 | 2.87 | 3.58 | 2.22 | 2.83 | 1.17 | 2.94 | 56.52 | 18.57 | 193.27 | 40.81 | 6.25 | 12.39 | 0.91 | 2.52 |
| 1992 | 2.78 | 2.26 | 2.32 | 2.54 | 1.68 | 2.08 | 44.82 | 16.00 | 91.34 | 35.78 | 5.22 | 16.61 | 8.42 | 2.59 |
| 1993 | 2.72 | 2.84 | 1.43 | 2.28 | 1.77 | 1.41 | 38.14 | 13.79 | 107.43 | 39.64 | 4.43 | 18.44 | 0.94 | 2.73 |
| 1994 | 2.71 | 2.47 | 2.18 | 3.07 | 0.83 | 4.14 | 23.36 | 9.48 | 84.97 | 37.03 | 3.03 | 9.48 | 0.24 | 1.94 |
| 1995 | 2.93 | 2.66 | 2.23 | 3.34 | 3.35 | 2.22 | 26.38 | 8.46 | 96.28 | 37.59 | 2.61 | 11.60 | 0.46 | 2.16 |
| 1996 | 2.63 | 2.05 | 1.91 | 1.84 | 0.38 | 0.77 | 23.60 | 8.67 | 81.18 | 39.78 | 4.60 | 8.70 | 1.68 | 3.91 |
| 1997 | 2.41 | 1.90 | 1.89 | 2.33 | 1.30 | 0.48 | 20.47 | 8.14 | 83.68 | 43.00 | 5.18 | 12.67 | 1.90 | 2.56 |
| 1998 | 1.59 | 1.54 | 1.24 | 0.93 | 0.33 | 0.69 | 10.94 | 7.13 | 85.06 | 47.84 | 5.09 | 10.45 | 1.55 | 2.81 |
| 1999 | 2.59 | 1.63 | 1.99 | 0.67 | 0.35 | 0.68 | 11.99 | 5.69 | 85.44 | 50.87 | 1.97 | 26.00 | 3.86 | 5.47 |
| 2000 | 2.29 | 1.00 | 3.10 | 0.68 | 0.19 | 0.60 | 10.98 | 4.05 | 53.46 | 51.19 | 2.56 | 17.53 | 2.34 | 3.36 |
| 2001 | 2.25 | 1.07 | 2.53 | 0.87 | 0.32 | 0.68 | 9.78 | 4.42 | 53.31 | 49.32 | 2.71 | 19.95 | 2.68 | 1.55 |
| 2002 | 1.31 | 1.14 | 3.70 | 1.49 | 0.54 | 0.27 | 6.81 | 6.10 | 37.93 | 37.53 | 1.54 | 6.19 | 2.49 | 0.93 |
| 2003 | 1.67 | 1.17 | 0.82 | 1.25 | 0.29 | 0.09 | 15.83 | 9.94 | 47.73 | 40.71 | 0.55 | 11.87 | 1.73 | 2.40 |
| 2004 | 1.28 | 1.16 | 0.93 | 0.51 | 0.18 | 0.22 | 12.44 | 9.42 | 40.06 | 32.37 | 3.03 | 14.25 | 2.03 | 2.42 |
| 2005 | 0.81 | 0.75 | 0.13 | 0.51 | 0.01 | 0.07 | 9.5 | 12.09 | 22.25 | 27.73 | 0.30 | 9.57 | 2.35 | 1.67 |
| 2006 | 1.53 | 0.88 | 0.47 | 0.91 | 0.05 | 0.03 | 19.78 | 12.97 | 13.99 | 18.57 | 0.31 | 10.48 | 3.47 | 1.16 |
| 2007 | 1.07 | 1.95 | 1.45 | 0.85 | 0.1 | 0.56 | 11.85 | 10.66 | 18.10 | 15.37 | 0.41 | 6.79 | 3.49 | 0.19 |
| 2008 | 1.27 | 2.95 | 1.69 | 0.8 | 0.01 | 0.1 | 13.21 | 10.13 | 18.80 | 13.83 | 1.58 | 3.84 | 3.65 | 0.08 |
| 2009* | 1.02 | 1.39 | 0.81 | 1.07 | 0.09 | 0.09 | 8.23 | 8.97 | 8.97 | 12.23 | 3.43 | 3.48 | 4.38 | 0.71 |

* Provisional

Table 7.10.3. Plaice in Divisions VIIf\&g. Lpue and effort for Irish and Belgian fleets in VIIf,g.

|  | IR-OTB-7G |  |  | IR-SCC-7G |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Effort (000 $\mathbf{h r}$ ) | LPUE (kg/h) | Landings (t) | Effort (000 hr) | LPUE (kg/h) |
| $\mathbf{1 9 9 5}$ | 94.23 | 63.56 | 1.48 | 9.55 | 6.43 | 1.49 |
| $\mathbf{1 9 9 6}$ | 133.66 | 60.04 | 2.23 | 14.20 | 9.73 | 1.46 |
| $\mathbf{1 9 9 7}$ | 119.84 | 65.10 | 1.84 | 38.79 | 16.13 | 2.40 |
| $\mathbf{1 9 9 8}$ | 96.72 | 72.30 | 1.34 | 21.38 | 14.94 | 1.43 |
| $\mathbf{1 9 9 9}$ | 60.05 | 51.66 | 1.16 | 10.40 | 8.01 | 1.30 |
| $\mathbf{2 0 0 0}$ | 28.78 | 60.60 | 0.47 | 11.40 | 9.90 | 1.15 |
| $\mathbf{2 0 0 1}$ | 23.82 | 69.43 | 0.34 | 10.93 | 16.33 | 0.67 |
| $\mathbf{2 0 0 2}$ | 42.30 | 77.69 | 0.54 | 16.42 | 20.86 | 0.79 |
| $\mathbf{2 0 0 3}$ | 26.35 | 86.79 | 0.30 | 13.80 | 20.91 | 0.66 |
| $\mathbf{2 0 0 4}$ | 26.62 | 96.99 | 0.27 | 5.04 | 19.38 | 0.26 |
| $\mathbf{2 0 0 5}$ | 22.78 | 124.40 | 0.18 | 6.46 | 14.81 | 0.44 |
| $\mathbf{2 0 0 6}$ | 24.58 | 118.36 | 0.21 | 5.10 | 14.79 | 0.34 |
| $\mathbf{2 0 0 7}$ | 30.38 | 135.41 | 0.22 | 4.76 | 15.81 | 0.30 |
| $\mathbf{2 0 0 8}$ | 39.17 | 125.81 | 0.31 | 8.38 | 11.65 | 0.72 |
| $\mathbf{2 0 0 9}$ | 43.09 | 135.18 | 0.32 | 7.87 | 8.15 | 0.97 |


|  | IR-TBB-7G |  |  | IR-GN-7G |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings (t) | Effort (000 hr) | LPUE (kg/h) | Landings (t) | Effort (000 hr) | LPUE (kg/h) |
| $\mathbf{1 9 9 5}$ | 37.92 | 20.78 | 1.83 | 0.90 | 4.33 | 0.21 |
| $\mathbf{1 9 9 6}$ | 53.02 | 26.76 | 1.98 | 1.35 | 5.51 | 0.24 |
| $\mathbf{1 9 9 7}$ | 94.59 | 28.25 | 3.35 | 1.17 | 1.51 | 0.78 |
| $\mathbf{1 9 9 8}$ | 122.13 | 35.25 | 3.46 |  | 0.00 |  |
| $\mathbf{1 9 9 9}$ | 25.80 | 40.87 | 0.63 | 0.48 | 5.47 | 0.09 |
| $\mathbf{2 0 0 0}$ | 12.62 | 37.03 | 0.34 | 2.54 | 7.03 | 0.36 |
| $\mathbf{2 0 0 1}$ | 4.80 | 39.71 | 0.12 | 0.30 | 4.46 | 0.07 |
| $\mathbf{2 0 0 2}$ | 7.08 | 31.62 | 0.22 | 0.36 | 5.86 | 0.06 |
| $\mathbf{2 0 0 3}$ | 9.37 | 49.26 | 0.19 | 0.20 | 10.97 | 0.02 |
| $\mathbf{2 0 0 4}$ | 6.17 | 54.86 | 0.11 | 0.33 | 12.05 | 0.03 |
| $\mathbf{2 0 0 5}$ | 9.49 | 49.65 | 0.19 | 0.12 | 10.89 | 0.01 |
| $\mathbf{2 0 0 6}$ | 14.40 | 60.35 | 0.24 | 0.09 | 7.76 | 0.01 |
| $\mathbf{2 0 0 7}$ | 20.35 | 54.85 | 0.37 | 0.32 | 8.83 | 0.04 |
| $\mathbf{2 0 0 8}$ | 14.18 | 37.22 | 0.38 | 0.01 | 13.13 | 0.00 |
| $\mathbf{2 0 0 9}$ | 6.96 | 37.92 | 0.18 | 0.07 | 12.61 | 0.01 |


|  |  |  |  |
| :---: | :---: | :---: | :---: |
|  | BELGIAN Beam Trawl VIlfg |  |  |
| Year | Landings (t) | Effort (000 $\mathbf{h r}$ ) PUE (kg/h |  |
| 1996 | 356.89 | 53.27 | 6.70 |
| 1997 | 474.71 | 57.36 | 8.28 |
| 1998 | 443.38 | 57.79 | 7.67 |
| 1999 | 410.22 | 55.11 | 7.44 |
| 2000 | 230.63 | 51.34 | 4.49 |
| 2001 | 274.84 | 54.90 | 5.01 |
| 2002 | 259.80 | 49.60 | 5.24 |
| 2003 | 215.95 | 62.73 | 3.44 |
| 2004 | 207.27 | 78.73 | 2.63 |
| 2005 | 153.73 | 64.50 | 2.38 |
| 2006 | 134.44 | 50.28 | 2.67 |
| 2007 | 139.39 | 45.72 | 3.05 |
| 2008 | 106.29 | 28.71 | 3.70 |
| 2009 | 215.99 | 30.84 | 7.00 |

Table 7.10.4. Plaice in Divisions VIIf\&g. Annual length distribution by fleet in 2009.


Table 7.10.5. Plaice in Divisions VIIf\&g. Catch numbers-at-age.

Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP
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| Table 1 | num | at age |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 |  |  |  |  |  |  |  |
| 1 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 2 | 989 | 851 | 877 |  |  |  |  |  |  |  |
| 3 | 426 | 903 | 673 |  |  |  |  |  |  |  |
| 4 | 411 | 291 | 638 |  |  |  |  |  |  |  |
| 5 | 105 | 136 | 72 |  |  |  |  |  |  |  |
| 6 | 72 | 76 | 70 |  |  |  |  |  |  |  |
| 7 | 37 | 47 | 34 |  |  |  |  |  |  |  |
| 8 | 59 | 23 | 8 |  |  |  |  |  |  |  |
| +gp | 75 | 98 | 46 |  |  |  |  |  |  |  |
| TOTALNUM | 2174 | 2425 | 2418 |  |  |  |  |  |  |  |
| TONSLAND | 757 | 875 | 863 |  |  |  |  |  |  |  |
| SOPCOF \% | 101 | 103 | 102 |  |  |  |  |  |  |  |
| AGE\YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 1921 | 822 | 300 | 750 | 704 | 1461 | 703 | 434 | 967 | 797 |
| 3 | 1207 | 2111 | 1180 | 560 | 918 | 2503 | 2595 | 1883 | 2099 | 3550 |
| 4 | 658 | 681 | 955 | 827 | 343 | 393 | 1332 | 1812 | 1568 | 1807 |
| 5 | 146 | 109 | 443 | 372 | 373 | 102 | 156 | 772 | 612 | 741 |
| 6 | 21 | 54 | 86 | 92 | 209 | 177 | 59 | 156 | 413 | 160 |
| 7 | 16 | 53 | 51 | 44 | 70 | 62 | 48 | 22 | 65 | 98 |
| 8 | 16 | 11 | 14 | 27 | 41 | 25 | 32 | 125 | 16 | 24 |
| +gp | 32 | 44 | 60 | 23 | 42 | 38 | 24 | 76 | 73 | 23 |
| TOTALNUM | 4017 | 3885 | 3089 | 2695 | 2700 | 4761 | 4949 | 5280 | 5813 | 7200 |
| TONSLAND | 1373 | 1377 | 1303 | 1146 | 1210 | 1752 | 1691 | 1901 | 2116 | 2151 |
| SOPCOF \% | 101 | 100 | 101 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| AGE\YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 164 | 279 | 800 | 1019 | 428 | 488 | 812 | 420 | 426 | 243 |
| 3 | 2078 | 1072 | 526 | 1179 | 936 | 572 | 734 | 1318 | 921 | 982 |
| 4 | 2427 | 1193 | 357 | 284 | 730 | 743 | 514 | 929 | 849 | 802 |
| 5 | 655 | 578 | 471 | 139 | 164 | 334 | 219 | 272 | 287 | 372 |
| 6 | 242 | 179 | 275 | 185 | 117 | 117 | 137 | 121 | 96 | 116 |
| 7 | 86 | 94 | 80 | 115 | 86 | 57 | 59 | 60 | 82 | 45 |
| 8 | 70 | 78 | 21 | 61 | 92 | 48 | 37 | 20 | 39 | 27 |
| +gp | 46 | 79 | 96 | 59 | 64 | 131 | 96 | 82 | 56 | 69 |
| TOTALNUM | 5768 | 3552 | 2626 | 3041 | 2617 | 2490 | 2608 | 3222 | 2756 | 2656 |
| TONSLAND | 2082 | 1501 | 1188 | 1114 | 1070 | 1028 | 952 | 1217 | 1067 | 968 |
| SOPCOF \% | 100 | 101 | 100 | 100 | 101 | 101 | 100 | 100 | 100 | 100 |
| AGE\YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 320 | 651 | 170 | 239 | 126 | 201 | 331 | 130 | 270 | 123 |
| 3 | 606 | 371 | 661 | 571 | 578 | 327 | 458 | 513 | 341 | 594 |
| 4 | 482 | 323 | 543 | 465 | 428 | 265 | 140 | 340 | 443 | 333 |
| 5 | 203 | 199 | 183 | 150 | 261 | 134 | 134 | 104 | 145 | 262 |
| 6 | 145 | 108 | 113 | 85 | 46 | 73 | 76 | 76 | 47 | 67 |
| 7 | 53 | 62 | 65 | 34 | 27 | 24 | 50 | 46 | 29 | 21 |
| 8 | 22 | 23 | 24 | 26 | 15 | 14 | 12 | 26 | 11 | 10 |
| +gp | 32 | 28 | 28 | 24 | 17 | 16 | 15 | 13 | 15 | 12 |
| TOTALNUM | 1862 | 1763 | 1786 | 1593 | 1498 | 1054 | 1217 | 1249 | 1300 | 1421 |
| TONSLAND | 718 | 714 | 642 | 594 | 510 | 386 | 404 | 410 | 437 | 463 |
| SOPCOF \% | 100 | 103 | 100 | 100 | 100 | 101 | 101 | 100 | 101 | 100 |

Table 7.10.6. Plaice in Divisions VIIf\&g. Catch weights-at-age.

Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP
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Table 2 Catch weights at age (kg)

| AGE\YEAR | 1977 | 1978 | 1979 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.078 | 0.194 | 0.076 |  |  |  |  |  |  |  |
| 2 | 0.205 | 0.258 | 0.203 |  |  |  |  |  |  |  |
| 3 | 0.323 | 0.323 | 0.325 |  |  |  |  |  |  |  |
| 4 | 0.43 | 0.389 | 0.44 |  |  |  |  |  |  |  |
| 5 | 0.528 | 0.457 | 0.55 |  |  |  |  |  |  |  |
| 6 | 0.615 | 0.525 | 0.652 |  |  |  |  |  |  |  |
| 7 | 0.693 | 0.595 | 0.749 |  |  |  |  |  |  |  |
| 8 | 0.76 | 0.666 | 0.839 |  |  |  |  |  |  |  |
| +gp | 0.8762 | 0.8435 | 1.0653 |  |  |  |  |  |  |  |
| SOPCOFAC | 1.0053 | 1.0265 | 1.0226 |  |  |  |  |  |  |  |
| AGE\YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 1 | 0.118 | 0.185 | 0.151 | 0.178 | 0.276 | 0.135 | 0 | 0.129 | 0.26 | 0.102 |
| 2 | 0.238 | 0.255 | 0.245 | 0.274 | 0.324 | 0.251 | 0.16 | 0.208 | 0.288 | 0.176 |
| 3 | 0.354 | 0.33 | 0.339 | 0.369 | 0.384 | 0.363 | 0.301 | 0.288 | 0.325 | 0.255 |
| 4 | 0.467 | 0.412 | 0.433 | 0.464 | 0.455 | 0.47 | 0.434 | 0.368 | 0.37 | 0.337 |
| 5 | 0.576 | 0.5 | 0.526 | 0.559 | 0.538 | 0.572 | 0.559 | 0.449 | 0.423 | 0.423 |
| 6 | 0.682 | 0.595 | 0.62 | 0.654 | 0.633 | 0.67 | 0.677 | 0.53 | 0.484 | 0.514 |
| 7 | 0.784 | 0.695 | 0.714 | 0.749 | 0.739 | 0.763 | 0.787 | 0.612 | 0.554 | 0.608 |
| 8 | 0.882 | 0.802 | 0.808 | 0.844 | 0.857 | 0.851 | 0.889 | 0.694 | 0.633 | 0.706 |
| +gp | 1.1812 | 1.1824 | 1.0948 | 1.1579 | 1.2661 | 1.0036 | 1.1033 | 0.8632 | 0.8887 | 0.9932 |
| SOPCOFAC | 1.0136 | 1.0043 | 1.0126 | 0.9997 | 1.0003 | 1.0048 | 0.9997 | 1.0034 | 1.0026 | 1.0007 |
| AGE\YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 0.24 | 0.2 | 0.148 | 0.172 | 0.145 | 0.22 | 0.222 | 0.181 | 0.188 | 0.096 |
| 2 | 0.27 | 0.26 | 0.257 | 0.247 | 0.24 | 0.264 | 0.26 | 0.248 | 0.248 | 0.188 |
| 3 | 0.309 | 0.327 | 0.362 | 0.326 | 0.331 | 0.319 | 0.309 | 0.318 | 0.316 | 0.279 |
| 4 | 0.358 | 0.4 | 0.464 | 0.407 | 0.42 | 0.382 | 0.368 | 0.392 | 0.39 | 0.369 |
| 5 | 0.416 | 0.481 | 0.563 | 0.492 | 0.506 | 0.456 | 0.438 | 0.469 | 0.471 | 0.457 |
| 6 | 0.483 | 0.567 | 0.658 | 0.58 | 0.589 | 0.539 | 0.519 | 0.55 | 0.559 | 0.545 |
| 7 | 0.56 | 0.661 | 0.75 | 0.671 | 0.67 | 0.632 | 0.609 | 0.634 | 0.655 | 0.631 |
| 8 | 0.646 | 0.761 | 0.839 | 0.765 | 0.747 | 0.735 | 0.711 | 0.723 | 0.757 | 0.716 |
| +gp | 0.9097 | 1.0465 | 1.0399 | 1.0061 | 0.9077 | 1.0351 | 0.9946 | 0.9972 | 1.1417 | 1.0022 |
| SOPCOFAC | 1.001 | 1.0115 | 1.0023 | 1.0031 | 1.0138 | 1.0104 | 1.0002 | 1.001 | 1.003 | 1.0021 |
| AGE\YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.145 | 0.248 | 0.132 | 0.183 | 0.14 | 0.176 | 0.257 | 0.163 | 0.262 | 0.272 |
| 2 | 0.226 | 0.299 | 0.202 | 0.24 | 0.204 | 0.229 | 0.261 | 0.212 | 0.272 | 0.297 |
| 3 | 0.309 | 0.354 | 0.278 | 0.305 | 0.273 | 0.293 | 0.284 | 0.267 | 0.294 | 0.297 |
| 4 | 0.394 | 0.414 | 0.358 | 0.38 | 0.347 | 0.366 | 0.326 | 0.33 | 0.321 | 0.308 |
| 5 | 0.481 | 0.478 | 0.444 | 0.463 | 0.426 | 0.449 | 0.386 | 0.399 | 0.371 | 0.345 |
| 6 | 0.57 | 0.547 | 0.535 | 0.556 | 0.511 | 0.542 | 0.465 | 0.476 | 0.446 | 0.41 |
| 7 | 0.661 | 0.62 | 0.631 | 0.657 | 0.602 | 0.645 | 0.563 | 0.56 | 0.544 | 0.502 |
| 8 | 0.753 | 0.697 | 0.733 | 0.767 | 0.697 | 0.757 | 0.68 | 0.651 | 0.666 | 0.621 |
| +gp | 1.0422 | 0.9739 | 1.0376 | 1.0235 | 0.9414 | 1.0386 | 0.9749 | 0.8497 | 0.9128 | 0.8924 |
| SOPCOFAC | 1.005 | 1.0277 | 1.001 | 1.0037 | 1.0015 | 1.0118 | 1.0086 | 1.0038 | 1.0236 | 1.0063 |

Table 7.10.7. Plaice in Divisions VIIf\&g. Stock weights-at-age. Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP

At 12/05/2010 18:34

Table 3 Stock weights at age (kg)

| YEAR | 1977 | 1978 | 1979 |
| ---: | ---: | ---: | ---: |
| 1 | 0.112 | 0.086 | 0.107 |
| 2 | 0.216 | 0.17 | 0.212 |
| 3 | 0.315 | 0.252 | 0.313 |
| 4 | 0.406 | 0.334 | 0.412 |
| 5 | 0.492 | 0.414 | 0.507 |
| 6 | 0.57 | 0.493 | 0.599 |
| 7 | 0.642 | 0.57 | 0.689 |
| 8 | 0.707 | 0.646 | 0.775 |
| + gp |  | 0.8389 | 0.8218 | 1.0148


| YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.109 | 0.082 | 0.096 | 0.103 | 0.256 | 0.075 | 0 | 0.089 | 0.249 | 0.066 |
| 2 | 0.217 | 0.167 | 0.192 | 0.206 | 0.298 | 0.193 | 0.087 | 0.168 | 0.273 | 0.139 |
| 3 | 0.322 | 0.257 | 0.288 | 0.307 | 0.352 | 0.307 | 0.232 | 0.248 | 0.305 | 0.215 |
| 4 | 0.426 | 0.35 | 0.383 | 0.408 | 0.418 | 0.417 | 0.369 | 0.328 | 0.346 | 0.295 |
| 5 | 0.528 | 0.447 | 0.479 | 0.507 | 0.495 | 0.521 | 0.498 | 0.408 | 0.395 | 0.38 |
| 6 | 0.628 | 0.548 | 0.574 | 0.606 | 0.584 | 0.621 | 0.619 | 0.489 | 0.453 | 0.468 |
| 7 | 0.727 | 0.653 | 0.668 | 0.704 | 0.685 | 0.717 | 0.733 | 0.571 | 0.518 | 0.56 |
| 8 | 0.823 | 0.762 | 0.763 | 0.801 | 0.797 | 0.808 | 0.839 | 0.653 | 0.593 | 0.657 |
| +gp | 1.1318 | 1.129 | 1.0492 | 1.1136 | 1.1897 | 0.9646 | 1.0635 | 0.8219 | 0.8373 | 0.938 |
| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 0.228 | 0.173 | 0.092 | 0.135 | 0.097 | 0.201 | 0.207 | 0.149 | 0.161 | 0.049 |
| 2 | 0.254 | 0.229 | 0.203 | 0.209 | 0.193 | 0.241 | 0.24 | 0.214 | 0.217 | 0.142 |
| 3 | 0.288 | 0.293 | 0.31 | 0.286 | 0.286 | 0.29 | 0.284 | 0.282 | 0.281 | 0.234 |
| 4 | 0.332 | 0.363 | 0.414 | 0.366 | 0.376 | 0.349 | 0.338 | 0.354 | 0.352 | 0.324 |
| 5 | 0.386 | 0.44 | 0.514 | 0.45 | 0.463 | 0.418 | 0.402 | 0.43 | 0.43 | 0.413 |
| 6 | 0.448 | 0.523 | 0.611 | 0.536 | 0.548 | 0.496 | 0.477 | 0.509 | 0.514 | 0.501 |
| 7 | 0.52 | 0.613 | 0.705 | 0.625 | 0.63 | 0.585 | 0.563 | 0.592 | 0.606 | 0.588 |
| 8 | 0.602 | 0.71 | 0.795 | 0.718 | 0.709 | 0.682 | 0.659 | 0.678 | 0.705 | 0.673 |
| +gp | 0.8537 | 0.987 | 1.0002 | 0.9544 | 0.8723 | 0.9712 | 0.9302 | 0.9476 | 1.0787 | 0.9622 |
| YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.105 | 0.224 | 0.099 | 0.158 | 0.11 | 0.153 | 0.262 | 0.142 | 0.154 | 0.154 |
| 2 | 0.185 | 0.273 | 0.167 | 0.21 | 0.171 | 0.201 | 0.257 | 0.187 | 0.199 | 0.199 |
| 3 | 0.268 | 0.326 | 0.239 | 0.271 | 0.238 | 0.26 | 0.27 | 0.239 | 0.252 | 0.253 |
| 4 | 0.352 | 0.384 | 0.317 | 0.341 | 0.309 | 0.328 | 0.303 | 0.298 | 0.305 | 0.299 |
| 5 | 0.438 | 0.446 | 0.401 | 0.42 | 0.386 | 0.406 | 0.353 | 0.364 | 0.343 | 0.323 |
| 6 | 0.525 | 0.512 | 0.489 | 0.508 | 0.468 | 0.494 | 0.423 | 0.437 | 0.405 | 0.374 |
| 7 | 0.615 | 0.583 | 0.583 | 0.605 | 0.556 | 0.592 | 0.512 | 0.517 | 0.492 | 0.453 |
| 8 | 0.707 | 0.658 | 0.682 | 0.711 | 0.649 | 0.7 | 0.619 | 0.604 | 0.602 | 0.558 |
| +gp | 0.9934 | 0.9283 | 0.9794 | 0.9582 | 0.8869 | 0.9709 | 0.8956 | 0.7964 | 0.8296 | 0.8068 |

Table 7.10.8. Plaice in Divisions VIIf\&g. Survey abundance indices (figures used in the assessment shown in bold).

IRGFS : Irish Groundfish Survey (IBTS 4th qtr VIIg)
20032008
110.790 .92

27

| 832 | 45 | 84 | 37 | 8 | 3 | 13 | 1 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 980 | 6 | 31 | 51 | 20 | 9 | 3 |  |
| 845 | 63 | 83 | 19 | 9 | 18 | 12 |  |
| 1046 | 105 | 80 | 22 | 22 | 9 | 8 |  |
| 1168 | 51 | 166 | 68 | 19 | 8 | 5 |  |
| 1139 | 113 | 106 | 72 | 100 | 21 | 16 |  |
| 1018 | 199 | 548 | 247 |  | 8 | 12 |  |

E+W B/T Survey
19902009 (Effort in Km towed, Numbers caught; all stations)
110.750 .85 (Revised 2008 - Indices automated 1995 on)

15
69.8616121564156
123.4184133652112
125.0848730713515
127.671201074425
120.821274020111
$\begin{array}{lllll}114.9 & 275 & 103 & 19 & 3 \\ 8\end{array}$
$118.6 \quad 2653423713$
114.92591174052
$114.9 \quad 27214454102$
$\begin{array}{llllll}118.6 & 181 & 94 & 34 & 23 & 8\end{array}$
$\begin{array}{llllll}118.6 & 403 & 75 & 37 & 8 & 7\end{array}$
118.625118519105

| 118.6 | 162 | 208 | 95 | 7 |
| :--- | :--- | :--- | :--- | :--- |

$\begin{array}{lllll}118.6 & 117 & 95 & 72 & 26 \\ 3\end{array}$
$\begin{array}{lllll}114.9 & 297 & 38 & 31 & 15\end{array}$
$\begin{array}{lllll}118.6 & 228 & 89 & 25 & 1013\end{array}$
118.610212141112
$118.6 \quad 17810956182$
$\begin{array}{llllll}118.6 & 167 & 257 & 57 & 19 & 6\end{array}$
$118.6 \quad 192 \quad 6693 \quad 2513$

Table 7.10.9. Plaice in Divisions VIIf\&g. Commercial tuning data available to the Working Group. (figures used in the assessment shown in bold).

## UK (E+W) BEAM TRAWL VIIF.

19902009 Thousands of hours, numbers in thousands.
1101
18

| 30.8 | 0.0 | $\mathbf{1 . 6}$ | $\mathbf{6 8 . 2}$ | $\mathbf{1 5 9 . 5}$ | $\mathbf{4 6 . 3}$ | $\mathbf{2 6 . 6}$ | $\mathbf{1 1 . 0}$ | 9.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 40.8 | 9.4 | 22.6 | 74.4 | 141.5 | 87.1 | 29.0 | 15.1 | 14.1 |
| 35.8 | 1.6 | 39.9 | 27.3 | 32.0 | 46.7 | 27.4 | 7.5 | 2.3 |
| 39.6 | 1.0 | 40.9 | 139.5 | 25.0 | 15.5 | 24.6 | 15.1 | 7.3 |
| 37 | 12.6 | 31.7 | 52.4 | 49.1 | 9.2 | 9.1 | 7.6 | 9.8 |
| 37.6 | 1.0 | 28.3 | 30.0 | 39.5 | 29.7 | 9.9 | 5.8 | 6.4 |
| 39.8 | 0.0 | 74.6 | 53.8 | 13.6 | 13.6 | 12.8 | 3.8 | 4.4 |
| 43 | 0.6 | 40.7 | 112.3 | 23.7 | 8.4 | 6.7 | 4.5 | 0.7 |
| 47.8 | 2.7 | 54.1 | 73.9 | 63.1 | 17.5 | 3.6 | 4.3 | 2.7 |
| 50.8 | 0.8 | 22.1 | 64.2 | 52.5 | 25.8 | 7.7 | 2.4 | 1.9 |
| 51.2 | 0.6 | 11.9 | 26.0 | 26.9 | 17.8 | 12.7 | 4.9 | 1.8 |
| 49.3 | 2.8 | 42.5 | 27.7 | 27.5 | 17.7 | 10.1 | 5.9 | 2.4 |
| 37.5 | 0.5 | 19.4 | 40.3 | 16.5 | 7.6 | 7.2 | 3.7 | 2.0 |
| 40.7 | 1.6 | 27.7 | 43.2 | 33.8 | 9.9 | 4.9 | 3.4 | 2.4 |
| 32.4 | 0.9 | 12.2 | 34.5 | 25.8 | 17.5 | 3.4 | 2.5 | 2.0 |
| 27.7 | 1.5 | 12.0 | 9.1 | 12.7 | 7.5 | 5.0 | 1.9 | 1.1 |
| 18.6 | 0.6 | 10.2 | 17.7 | 4.5 | 4.4 | 3.0 | 1.6 | 0.4 |
| 15.4 | 0.5 | 9.3 | 24.6 | 12.0 | 3.2 | 2.0 | 1.4 | 0.6 |
| 13.8 | 0.2 | 10.8 | 16.1 | 18.1 | 5.2 | 1.9 | 1.4 | 0.9 |
| 12.2 | 0.3 | 10.4 | 30.1 | 15.2 | 10.6 | 3.0 | 1.0 | 0.6 |

## UK(E+W) OTTER TRAWL VIIF

19892009 Thousands of hours, numbers in thousands.
1101
18

| 17.6 | 0.8 | 91.2 | $\mathbf{2 5 6 . 0}$ | $\mathbf{6 2 . 0}$ | $\mathbf{2 3 . 1}$ | $\mathbf{7 . 4}$ | $\mathbf{5 . 1}$ | 0.4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 22.6 | 0.1 | 6.4 | 97.0 | 129.1 | 34.2 | 13.3 | 4.1 | 4.4 |
| 18.6 | 5.2 | 13.6 | 46.9 | 78.8 | 36.9 | 16.5 | 4.4 | 5.0 |
| 16.0 | 3.6 | 68.2 | 14.6 | 12.5 | 18.5 | 8.5 | 1.4 | 0.4 |
| 13.8 | 1.3 | 25.3 | 42.1 | 8.8 | 3.9 | 6.3 | 4.1 | 2.7 |
| 9.5 | 4.2 | 11.7 | 20.5 | 15.1 | 2.7 | 3.1 | 1.4 | 1.7 |
| 8.5 | 5.1 | 37.8 | 18.2 | 14.5 | 5.5 | 1.6 | 0.8 | 0.7 |
| 8.7 | 0.0 | 35.8 | 20.6 | 4.3 | 3.4 | 2.5 | 1.0 | 1.1 |
| 8.1 | 0.4 | 16.5 | 33.7 | 5.5 | 1.2 | 0.7 | 0.4 | 0.1 |
| 7.1 | 0.4 | 7.8 | 11.0 | 8.6 | 2.0 | 0.5 | 0.7 | 0.2 |
| 5.7 | 1.0 | 8.3 | 12.2 | 7.9 | 3.8 | 0.9 | 0.2 | 0.1 |
| 4.1 | 0.5 | 9.3 | 11.4 | 6.5 | 2.5 | 1.3 | 0.4 | 0.1 |
| 4.4 | 1.4 | 11.1 | 4.9 | 4.0 | 2.4 | 1.3 | 0.6 | 0.2 |
| 6.1 | 0.0 | 4.4 | 8.3 | 2.9 | 1.5 | 1.1 | 0.5 | 0.2 |
| 9.9 | 0.6 | 11.9 | 16.2 | 9.3 | 2.1 | 1.3 | 0.9 | 0.6 |
| 9.4 | 0.3 | 4.3 | 14.3 | 10.4 | 5.8 | 0.9 | 0.5 | 0.3 |
| 12.1 | 1.5 | 10.0 | 5.4 | 5.5 | 2.8 | 1.5 | 0.5 | 0.3 |
| 13.0 | 0.7 | 12.8 | 23.3 | 6.8 | 6.4 | 4.5 | 2.3 | 0.6 |
| 10.6 | 0.2 | 5.2 | 14.8 | 7.4 | 2.2 | 1.4 | 1.0 | 0.5 |
| 10.1 | 0.3 | 5.8 | 16.5 | 8.2 | 2.4 | 1.6 | 1.1 | 0.6 |
| 9.0 | 0.2 | 5.6 | 7.8 | 7.3 | 2.3 | 0.9 | 0.5 | 0.3 |

Table 7.10.10. Plaice in Divisions VIIf\&g. XSA Diagnostics.

```
Lowestoft VPA Version 3.1
    15/05/2010 10:55
Extended Survivors Analysis
CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP
CPUE data from file P7FTUN3.DAT
Catch data for }33\mathrm{ years. }1977\mathrm{ to 2009. Ages }1\mathrm{ to 9.
\begin{tabular}{lrrrrrrrr} 
Fleet & \multicolumn{2}{c}{ First Last } & First & \multicolumn{2}{c}{ Last } & \multicolumn{2}{c}{ Alpha } & \multicolumn{2}{c}{ Beta } \\
& year & year & age & & age & & & \\
JK(E+W) BEAM TRAWL V & 1990 & 2009 & & 4 & 8 & 0 & 1 \\
UK(EW)OTTER TRAWL VI & 1989 & 2009 & 4 & 8 & 0 & 1 \\
EW BT Survey & 1990 & 2009 & 1 & 5 & 0.75 & 0.85
\end{tabular}
Time series weights :
    Tapered time weighting not applied
Catchability analysis:
    Catchability dependent on stock size for ages < 6
    Regression type = C
    Minimum of }5\mathrm{ points used for regression
    Survivor estimates shrunk to the population mean for ages < 6
    Catchability independent of age for ages >= 7
Terminal population estimation :
    Survivor estimates shrunk towards the mean F
    of the final }5\mathrm{ years or the 4 oldest ages.
    S.E. of the mean to which the estimates are shrunk = 2.500
    Minimum standard error for population
    estimates derived from each fleet = . }50
    Prior weighting not applied
Tuning converged after 71 iterations
```

Regression weights

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.

| Fishing mortalities <br> Age |  |  |  |  |  |  |  |  | 2000 | 2001 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |

1
XSA population numbers (Thousands)

| AGE |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 2000 | $3.23 \mathrm{E}+03$ | $1.95 \mathrm{E}+03$ | $1.53 \mathrm{E}+03$ | $1.05 \mathrm{E}+03$ | $5.33 \mathrm{E}+02$ | $3.07 \mathrm{E}+02$ | $1.19 \mathrm{E}+02$ | $3.40 \mathrm{E}+01$ |
| 2001 | $2.28 \mathrm{E}+03$ | $2.86 \mathrm{E}+03$ | $1.43 \mathrm{E}+03$ | $7.86 \mathrm{E}+02$ | $4.76 \mathrm{E}+02$ | $2.81 \mathrm{E}+02$ | $1.36 \mathrm{E}+02$ | $5.56 \mathrm{E}+01$ |
| 2002 | $2.04 \mathrm{E}+03$ | $2.02 \mathrm{E}+03$ | $1.93 \mathrm{E}+03$ | $9.18 \mathrm{E}+02$ | $3.93 \mathrm{E}+02$ | $2.36 \mathrm{E}+02$ | $1.48 \mathrm{E}+02$ | $6.24 \mathrm{E}+01$ |
| 2003 | $1.26 \mathrm{E}+03$ | $1.81 \mathrm{E}+03$ | $1.63 \mathrm{E}+03$ | $1.09 \mathrm{E}+03$ | 3.02E+02 | $1.77 \mathrm{E}+02$ | $1.03 \mathrm{E}+02$ | $6.97 \mathrm{E}+01$ |
| 2004 | $2.03 \mathrm{E}+03$ | $1.12 \mathrm{E}+03$ | $1.38 \mathrm{E}+03$ | $9.12 \mathrm{E}+02$ | $5.27 \mathrm{E}+02$ | $1.27 \mathrm{E}+02$ | 7.65E+01 | $5.88 \mathrm{E}+01$ |
| 2005 | $2.81 \mathrm{E}+03$ | $1.80 \mathrm{E}+03$ | $8.73 \mathrm{E}+02$ | $6.78 \mathrm{E}+02$ | $4.06 \mathrm{E}+02$ | $2.21 \mathrm{E}+02$ | $6.98 \mathrm{E}+01$ | $4.27 \mathrm{E}+01$ |
| 2006 | $2.16 \mathrm{E}+03$ | $2.50 \mathrm{E}+03$ | $1.41 \mathrm{E}+03$ | $4.66 \mathrm{E}+02$ | $3.52 \mathrm{E}+02$ | $2.34 \mathrm{E}+02$ | $1.28 \mathrm{E}+02$ | $3.90 \mathrm{E}+01$ |
| 2007 | $3.54 \mathrm{E}+03$ | $1.91 \mathrm{E}+03$ | $1.90 \mathrm{E}+03$ | $8.19 \mathrm{E}+02$ | $2.81 \mathrm{E}+02$ | $1.86 \mathrm{E}+02$ | $1.36 \mathrm{E}+02$ | $6.58 \mathrm{E}+01$ |
| 2008 | $2.19 \mathrm{E}+03$ | $3.14 \mathrm{E}+03$ | $1.57 \mathrm{E}+03$ | $1.20 \mathrm{E}+03$ | $4.06 \mathrm{E}+02$ | $1.52 \mathrm{E}+02$ | $9.35 \mathrm{E}+01$ | $7.74 \mathrm{E}+01$ |
| 2009 | $3.02 \mathrm{E}+03$ | $1.94 \mathrm{E}+03$ | $2.53 \mathrm{E}+03$ | $1.07 \mathrm{E}+03$ | . 5 | 2.24 | $9.01 \mathrm{E}+01$ | $5.59 \mathrm{E}+01$ |

Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+00 \quad 2.68 \mathrm{E}+03 \quad 1.61 \mathrm{E}+03 \quad 1.68 \mathrm{E}+03 \quad 6.39 \mathrm{E}+02 \quad 3.29 \mathrm{E}+02 \quad 1.35 \mathrm{E}+02 \quad 6.04 \mathrm{E}+01$

Taper weighted geometric mean of the VPA populations:
$3.75 \mathrm{E}+03 \quad 3.36 \mathrm{E}+03 \quad 2.47 \mathrm{E}+03 \quad 1.30 \mathrm{E}+03 \quad 5.64 \mathrm{E}+02 \quad 2.70 \mathrm{E}+02 \quad 1.34 \mathrm{E}+02 \quad 6.94 \mathrm{E}+01$

Standard error of the weighted Log(VPA populations) :

| 0.5609 | 0.5596 | 0.5708 | 0.5682 | 0.5561 | 0.5304 | 0.5704 | 0.7299 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g :XSA Diagnostics.

Log catchability residuals.
Fleet : UK(E+W) BEAM TRAWL V

| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 0.09 | 0.29 | 0.15 | 0.1 | 0.04 | 0.01 | -0.63 | -0.6 | -0.06 | -0.01 |
|  | 5 | 0.02 | 0.06 | 0.1 | -0.12 | -0.17 | 0.21 | -0.17 | -0.4 | -0.16 | -0.15 |
|  | 6 | 0.86 | 0.39 | 0.11 | 0.36 | -0.21 | 0.37 | 0.28 | -0.4 | -0.74 | -0.33 |
|  | 7 | -0.07 | 0.91 | -0.3 | -0.03 | -0.14 | -0.11 | 0.37 | -0.05 | -0.01 | -0.31 |
|  | 8 | 0.16 | 0.33 | -0.17 | 0.05 | 0.11 | 0.21 | 0.07 | -0.19 | 0.26 | 0.33 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 4 | -0.25 | 0.05 | -0.13 | 0.03 | 0.2 | 0.05 | -0.09 | 0.26 | 0.22 | 0.27 |
|  | 5 | -0.14 | 0.01 | -0.13 | 0.25 | 0.23 | -0.06 | 0.01 | 0.15 | 0.17 | 0.29 |
|  | 6 | 0.01 | -0.17 | 0.05 | -0.15 | -0.07 | -0.1 | -0.27 | -0.19 | 0 | 0.18 |
|  | 7 | 0 | 0.1 | -0.24 | -0.08 | 0.15 | 0.12 | -0.22 | -0.27 | 0.19 | -0.04 |
|  | 8 | 0.46 | 0.06 | -0.09 | -0.01 | 0.13 | 0.05 | -0.48 | -0.34 | -0.17 | -0.11 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: |
| Mean $\log q$ | -6.7505 | -6.772 | -6.772 |
| S.E(Log q) | 0.3493 | 0.2779 | 0.2369 |

Regression statistics :

Ages with q dependent on year class strength

|  |
| :---: |
| ge |


| 4 | 0.73 | 2.233 | 6.91 | 0.79 | 20 | 0.26 | -6.85 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 | 0.67 | 3.384 | 6.64 | 0.86 | 20 | 0.19 | -6.81 |

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |  | Mean Q |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
|  | 6 | 0.72 | 2.372 | 6.43 | 0.8 | 20 | 0.23 | -6.75 |  |  |
|  | 7 | 1.07 | -0.547 | 6.9 | 0.76 | 20 | 0.3 | -6.77 |  |  |
|  | 8 | 0.92 | 1.188 | 6.55 | 0.92 | 20 | 0.21 | -6.74 |  |  |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.

Fleet : UK(EW)OTTER TRAWL VI

Age
1989
No data for this fleet at this age
No data for this fleet at this age
No data for this fleet at this age
-0.02
0.05
-0.25
-0.07
-0.06

| Age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

No data for this fleet at this age
No data for this fleet at this age
No data for this fleet at this age

|  | 4 | 0.17 | 0.41 | -0.1 | -0.06 | 0.08 | 0.28 | -0.54 | -0.59 | -0.24 | 0.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5 | 0.13 | 0.12 | 0.1 | -0.35 | -0.09 | 0.12 | -0.07 | -0.63 | -0.34 | 0.09 |
|  | 6 | 0.56 | 0.7 | -0.17 | 0.14 | 0.16 | 0.12 | 0.26 | -0.9 | -0.72 | -0.2 |
|  | 7 | -0.52 | 0.69 | -0.95 | -0.06 | -0.24 | -0.38 | 0.78 | -0.57 | 0.31 | -0.38 |
|  | 8 | -0.05 | 0.3 | -0.89 | 0.33 | -0.06 | -0.29 | 0.43 | -0.24 | -0.22 | -0.2 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 No data for this fleet at this age <br> 2 No data for this fleet at this age <br> 3 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 0.46 | 0.27 | -0.23 | -0.02 | 0.32 | -0.15 | 0.29 | 0.02 | -0.27 | -0.19 |
|  | 5 | 0.28 | 0.33 | -0.03 | 0.14 | 0.37 | -0.19 | 0.53 | 0.12 | -0.16 | -0.55 |
|  | 6 | 0.34 | 0.28 | 0.05 | 0.02 | -0.07 | -0.39 | 0.58 | -0.09 | 0.22 | -0.64 |
|  | 7 | 0.25 | 0.46 | -0.37 | 0.23 | 0 | -0.16 | 0.73 | -0.01 | 0.49 | -0.2 |
|  | 8 | 0.32 | 0.21 | -0.25 | 0.24 | -0.31 | -0.2 | 0.51 | 0.07 | -0.04 | -0.27 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: |
| Mean Log q | -6.8357 | -6.9957 | -6.9957 |
| S.E(Log q) | 0.4209 | 0.4654 | 0.3288 |

Regression statistics :
Ages with q dependent on year class strength

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |  | Mean Log q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| 4 | 0.8 | 1.605 | 6.75 | 0.77 | 21 | 0.3 | -6.66 |  |  |  |
| 5 | 0.73 | 1.877 | 6.7 | 0.72 | 21 | 0.3 | -6.82 |  |  |  |

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |  | Mean Q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| 6 | 0.84 | 0.961 | 6.64 | 0.65 | 21 | 0.35 | -6.84 |  |  |  |
| 7 | 1.4 | -1.588 | 7.78 | 0.46 | 21 | 0.63 | -7 |  |  |  |
|  | 0.95 | 0.51 | 6.89 | 0.84 | 21 | 0.32 | -7.03 |  |  |  |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.

Fleet: UK (BT-Q3) Survey

| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.52 | 1.2 | 0.5 | -0.98 | -1.14 | -0.31 | -0.09 | 0.06 | 0.53 | -0.05 |
|  | 2 | 0.75 | -0.74 | 0.23 | -0.49 | -0.84 | -0.4 | 0.25 | -0.28 | 0.03 | 0.06 |
|  | 3 | 0.57 | 0.74 | -1.33 | -0.33 | -1.26 | -1.04 | -0.55 | -0.43 | 0.16 | -0.18 |
|  | 4 | 0.62 | 0.53 | -1.17 | -2.48 | -0.06 | -1.79 | -3.58 | -1.03 | -0.19 | 1.47 |
|  | 5 | -0.06 | -0.25 | 0.36 | -0.21 | -1.33 | 0.35 | -0.48 | -0.56 | -0.79 | 0.28 |
|  | 6 | No data fo | fleet a | s age |  |  |  |  |  |  |  |
|  | 7 | No data fo | fleet a | s age |  |  |  |  |  |  |  |
|  | 8 | No data for | fleet a | age |  |  |  |  |  |  |  |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 | 0.65 | 0.36 | -0.13 | -0.09 | 0.74 | 0.02 | -0.81 | -0.55 | -0.16 | -0.29 |
|  | 2 | -0.06 | 0.3 | 0.63 | 0.17 | -0.05 | 0.11 | 0.03 | 0.17 | 0.35 | -0.23 |
|  | 3 | 0.21 | -0.89 | 1.19 | 0.98 | 0.16 | 0.19 | 0.3 | 0.34 | 0.47 | 0.7 |
|  | 4 | -0.17 | 0.35 | 0.16 | 1.6 | 1.06 | 0.45 | 0.79 | 1.27 | 0.86 | 1.29 |
|  | 5 | 0.31 | 0.15 | 0.73 | 0.25 | -0.27 | 1.09 | -0.44 | -0.23 | 0.4 | 0.71 |
|  | 6 | No data fo | fleet a | s age |  |  |  |  |  |  |  |
|  | 7 | No data fo | fleet a | s age |  |  |  |  |  |  |  |
|  |  | No data for | fleet a | age |  |  |  |  |  |  |  |

Regression statistics :
Ages with q dependent on year class strength
Age Slope t-value Intercept RSquare No Pts Reg s.e Mean Log q

| 1 | 1.36 | -0.899 | 6.94 | 0.26 | 20 | 0.61 | -7.2 |
| ---: | ---: | ---: | ---: | ---: | :--- | :--- | ---: |
| 2 | 0.77 | 0.854 | 7.67 | 0.44 | 20 | 0.42 | -7.63 |
| 3 | 1.39 | -0.88 | 8.42 | 0.22 | 20 | 0.75 | -8.19 |
| 4 | 1.6 | -0.896 | 10.1 | 0.11 | 20 | 1.42 | -8.95 |
| 5 | 0.93 | 0.237 | 8.73 | 0.37 | 20 | 0.59 | -8.92 |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.

Terminal year survivor and $F$ summaries :
Age 1 Catchability dependent on age and year class strength

| Year class $=2008$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated <br> F |
|  | Survivors | s.e | s.e | Ratio |  | Weights |  |
| UK(E+W) BEAM TRAWL V | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK(EW)OTTER TRAWL VI | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (BT-Q3) Survey | 2010 | 0.632 | 0 | 0 | 1 | 0.44 | 0 |
| P shrinkage mean | 3357 | 0.56 |  |  |  | 0.56 | 0 |
| F shrinkage mean | 0 | 2.5 |  |  |  | 0 | 0 |

Weighted prediction :

| Survivors |  | Int | Ext | $N$ |  | Var | $F$ |
| :--- | ---: | ---: | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  | Ratio |  |  |
|  | 2679 | 0.42 | 0.38 |  | 2 | 0.917 | 0 |

Age 2 Catchability dependent on age and year class strength Year class $=2007$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK(E+W) BEAM TRAWL V | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK(EW)OTTER TRAWL VI | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (BT-Q3) Survey | 1312 | 0.393 | 0.038 | 0.1 | 2 | 0.652 | 0.084 |
| P shrinkage mean | 2469 | 0.57 |  |  |  | 0.331 | 0.046 |
| F shrinkage mean | 943 | 2.5 |  |  |  | 0.017 | 0.116 |

Weighted prediction :

| Survivors |  | Int | Ext | N |  | Var |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  | Ratio |  |
|  | 1608 | 0.32 | 0.21 | 4 | 0.661 | 0.1 |

Age 3 Catchability dependent on age and year class strength Year class $=2006$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK(E+W) BEAM TRAWL V | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK(EW)OTTER TRAWL VI | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (BT-Q3) Survey | 1963 | 0.352 | 0.34 | 0.97 | 3 | 0.633 | 0.251 |
| P shrinkage mean | 1303 | 0.57 |  |  |  | 0.349 | 0.357 |
| F shrinkage mean | 1054 | 2.5 |  |  |  | 0.018 | 0.426 |

Weighted prediction :

Survivors
at end of year

| Int | Ext | N |  | Var | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| s.e | s.e |  |  | Ratio |  |
| 0.3 | 0.23 |  | 5 | 0.764 | 0.3 |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.

Age 4 Catchability dependent on age and year class strength
Year class $=2005$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | $N$ | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK(E+W) BEAM TRAWL V | 836 | 0.5 | 0 | 0 | 1 | 0.208 | 0.318 |
| UK(EW)OTTER TRAWL VI | 527 | 0.5 | 0 | 0 | 1 | 0.208 | 0.467 |
| UK (BT-Q3) Survey | 681 | 0.347 | 0.34 | 0.98 | 4 | 0.323 | 0.379 |
| P shrinkage mean | 564 | 0.56 |  |  |  | 0.25 | 0.442 |
| F shrinkage mean | 438 | 2.5 |  |  |  | 0.012 | 0.54 |

Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |  |
| :--- | :---: | :---: | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  | Ratio |  |  |
|  | 639 | 0.23 | 0.14 |  | 8 | 0.615 | 0.4 |

Age 5 Catchability dependent on age and year class strength
Year class $=2004$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK(E+W) BEAM TRAWL V | 429 | 0.364 | 0.031 | 0.08 | 2 | 0.254 | 0.454 |
| UK(EW)OTTER TRAWL VI | 212 | 0.364 | 0.133 | 0.37 | 2 | 0.254 | 0.772 |
| UK (BT-Q3) Survey | 489 | 0.331 | 0.164 | 0.5 | 5 | 0.236 | 0.408 |
| P shrinkage mean | 270 | 0.53 |  |  |  | 0.245 | 0.649 |
| F shrinkage mean | 347 | 2.5 |  |  |  | 0.011 | 0.537 |

Weighted prediction :

| Survivors |  | Int | Ext | N | Var | F |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  | Ratio |  |  |
|  | 329 | 0.2 | 0.12 |  | 11 | 0.603 | 0.6 |

Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=2003$

| Fleet | Estimated  <br> Survivors Int | s.e | sxt | Var | N | Scaled | Estimated |
| :--- | ---: | :--- | :---: | :---: | :---: | :---: | ---: |
|  | 163 | 0.312 | 0.022 | 0.07 | 3 | 0.403 | 0.328 |
| UK(E+W) BEAM TRAWL V | 93 | 0.312 | 0.197 | 0.63 | 3 | 0.403 | 0.518 |
| UK(EW)OTTER TRAWL VI | 204 | 0.341 | 0.133 | 0.39 | 5 | 0.181 | 0.271 |
| UK (BT-Q3) Survey |  |  |  |  |  |  |  |
| F shrinkage mean | 108 | 2.5 |  |  |  | 0.012 | 0.462 |

Weighted prediction :

| Survivors |  | Int | Ext | N |  | Var | F |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  | Ratio |  |  |
|  | 135 | 0.19 | 0.11 |  | 12 | 0.602 | 0.4 |

Table 7.10.10. (cont.) Plaice in Divisions VIIf\&g. XSA Diagnostics.
Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=2002$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK(E+W) BEAM TRAWL V | 60 | 0.276 | 0.042 | 0.15 | 4 | 0.44 | 0.28 |
| UK(EW)OTTER TRAWL VI | 62 | 0.276 | 0.119 | 0.43 | 4 | 0.44 | 0.272 |
| UK (BT-Q3) Survey | 57 | 0.338 | 0.123 | 0.36 | 5 | 0.111 | 0.295 |
| F shrinkage mean | 33 | 2.5 |  |  |  | 0.01 | 0.463 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 60 | 0.18 | 0.05 | 14 | 0.28 | 0.3 |  |  |

Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 7

Year class $=2001$

| Fleet | Estimated  <br> Survivors Int | s.e | s.e | Var | Ratio |  | Scaled |  | Weights | Estimated |
| :--- | ---: | :--- | :---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: |
|  | 40 | 0.267 | 0.072 | 0.27 | 5 | 0.467 | 0.209 |  |  |  |
| UK(E+W) BEAM TRAWL V | 42 | 0.267 | 0.175 | 0.66 | 5 | 0.467 | 0.197 |  |  |  |
| UK(EW)OTTER TRAWL VI | 34 | 0.356 | 0.154 | 0.43 | 5 | 0.057 | 0.237 |  |  |  |
| UK (BT-Q3) Survey |  |  |  |  |  | 0.009 | 0.407 |  |  |  |
| F shrinkage mean | 18 | 2.5 |  |  |  | 0.0 |  |  |  |  |


| Survivors | Int | Ext | N | Var | F |  |  |
| :--- | ---: | ---: | :--- | :--- | :--- | :--- | :--- |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
|  | 40 | 0.18 | 0.07 |  | 16 | 0.41 | 0.2 |

Table 7.10.11. Plaice in Divisions VIIf\&g. Fishing Mortalities.
Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP
At 15/05/2010 $10: 56$ At 15/05/2010 10:56

Terminal Fs derived using XSA (With F shrinkage)
Table 8 Fishing mortality ( $F$ ) at age
AGE\YEAR 197719781979

| 1 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: |
| 2 | 0.3501 | 0.3347 | 0.2376 |
| 3 | 0.5445 | 0.5649 | 0.4377 |
| 4 | 0.8499 | 0.8152 | 0.9283 |
| 5 | 0.4485 | 0.6925 | 0.4329 |
| 6 | 0.6853 | 0.6196 | 0.8666 |
| 7 | 0.6003 | 1.2891 | 0.5671 |
| 8 | 0.6493 | 0.8593 | 0.7025 |
| +gp | 0.6493 | 0.8593 | 0.7025 |
| FBAR 3-6 | 0.632 | 0.673 | 0.6664 |


|  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| AGE YYEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.3388 | 0.1954 | 0.1927 | 0.2921 | 0.0958 | 0.1883 | 0.1126 | 0.0654 | 0.1011 | 0.1403 |
| 3 | 0.5382 | 0.6923 | 0.4297 | 0.5929 | 0.6314 | 0.5161 | 0.536 | 0.4461 | 0.46 | 0.5801 |
| 4 | 0.9296 | 0.6043 | 0.7117 | 0.5519 | 0.8183 | 0.5532 | 0.5197 | 0.8173 | 0.7503 | 0.8369 |
| 5 | 0.5023 | 0.3374 | 0.9385 | 0.6081 | 0.4693 | 0.5529 | 0.4014 | 0.5894 | 0.6567 | 0.9053 |
| 6 | 0.1962 | 0.3176 | 0.4417 | 0.4531 | 0.7554 | 0.3863 | 0.6568 | 0.8141 | 0.6621 | 0.3198 |
| 7 | 0.439 | 0.9595 | 0.5076 | 0.3862 | 0.6768 | 0.4739 | 0.1557 | 0.4949 | 0.8929 | 0.2893 |
| 8 | 0.5191 | 0.5573 | 0.6532 | 0.502 | 0.6836 | 0.4937 | 0.4351 | 0.6825 | 0.7447 | 0.9174 |
| +gp | 0.5191 | 0.5573 | 0.6532 | 0.502 | 0.6836 | 0.4937 | 0.4351 | 0.6825 | 0.7447 | 0.9174 |
| FBAR 3-6 | 0.5416 | 0.4879 | 0.6304 | 0.5515 | 0.6686 | 0.5021 | 0.5285 | 0.6667 | 0.6323 | 0.6605 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE\YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.0667 | 0.166 | 0.2237 | 0.3162 | 0.1947 | 0.16 | 0.2051 | 0.1408 | 0.1658 | 0.1394 |
| 3 | 0.585 | 0.7096 | 0.4842 | 0.5388 | 0.4864 | 0.3914 | 0.3486 | 0.5391 | 0.4683 | 0.6321 |
| 4 | 0.9314 | 0.7229 | 0.491 | 0.4768 | 0.69 | 0.8227 | 0.6642 | 0.9042 | 0.7316 | 0.8827 |
| 5 | 0.7659 | 0.5327 | 0.6387 | 0.3261 | 0.507 | 0.7187 | 0.5521 | 0.8269 | 0.7175 | 0.7615 |
| 6 | 0.7814 | 0.4379 | 0.4738 | 0.504 | 0.4556 | 0.7575 | 0.6664 | 0.6144 | 0.717 | 0.6529 |
| 7 | 0.2592 | 0.7314 | 0.3242 | 0.3366 | 0.4208 | 0.3813 | 1.0333 | 0.6306 | 1.0451 | 0.8104 |
| 8 | 0.3149 | 0.3604 | 0.3168 | 0.3993 | 0.4477 | 0.3993 | 0.4151 | 1.1773 | 1.0327 | 1.1861 |
| +gp | 0.3149 | 0.3604 | 0.3168 | 0.3993 | 0.4477 | 0.3993 | 0.4151 | 1.1773 | 1.0327 | 1.1861 |
| FBAR 3-6 | 0.7659 | 0.6008 | 0.5219 | 0.4614 | 0.5348 | 0.6726 | 0.5578 | 0.7211 | 0.6586 | 0.7323 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE\YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.1915 | 0.2763 | 0.0932 | 0.1512 | 0.1278 | 0.1262 | 0.152 | 0.075 | 0.0957 | 0.0693 |
| 3 | 0.5458 | 0.3224 | 0.4525 | 0.4634 | 0.5894 | 0.508 | 0.423 | 0.3379 | 0.2615 | 0.287 |
| 4 | 0.6688 | 0.5722 | 0.9904 | 0.6046 | 0.6899 | 0.536 | 0.3835 | 0.5815 | 0.4961 | 0.3991 |
| 5 | 0.5192 | 0.5844 | 0.6807 | 0.7453 | 0.7482 | 0.4303 | 0.5168 | 0.4994 | 0.4763 | 0.5589 |
| 6 | 0.6967 | 0.524 | 0.7111 | 0.7167 | 0.4802 | 0.4296 | 0.4216 | 0.569 | 0.3996 | 0.3853 |
| 7 | 0.6411 | 0.6564 | 0.6301 | 0.4368 | 0.463 | 0.4613 | 0.5423 | 0.4444 | 0.3944 | 0.2796 |
| 8 | 1.1314 | 0.5656 | 0.5304 | 0.5065 | 0.3209 | 0.4165 | 0.4072 | 0.5496 | 0.1635 | 0.206 |
| +gp | 1.1314 | 0.5656 | 0.5304 | 0.5065 | 0.3209 | 0.4165 | 0.4072 | 0.5496 | 0.1635 | 0.206 |
| FBAR 3-6 | 0.6076 | 0.5008 | 0.7087 | 0.6325 | 0.6269 | 0.476 | 0.4362 | 0.4969 | 0.4084 | 0.4076 |

FBAR 07-09
0
0.08
0.2955
0.4922
0.5115
0.4513
0.3728
0.3064

Table 7.10.12. Plaice in Divisions VIIf\&g. Population numbers.

Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP
At 15/05/2010 10:56

Terminal Fs derived using XSA (With F shrinkage)

| Table 10 | Stock nu | er at age | tart of ye |  | Numbers*10**-3 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE\YEAR | 1977 | 1978 | 1979 |  |  |  |  |  |  |  |  |  |  |
| 1 | 3582 | 4964 | 8004 |  |  |  |  |  |  |  |  |  |  |
| 2 | 3555 | 3177 | 4403 |  |  |  |  |  |  |  |  |  |  |
| 3 | 1077 | 2222 | 2016 |  |  |  |  |  |  |  |  |  |  |
| 4 | 762 | 554 | 1120 |  |  |  |  |  |  |  |  |  |  |
| 5 | 309 | 289 | 218 |  |  |  |  |  |  |  |  |  |  |
| 6 | 154 | 175 | 128 |  |  |  |  |  |  |  |  |  |  |
| 7 | 87 | 69 | 83 |  |  |  |  |  |  |  |  |  |  |
| 8 | 131 | 42 | 17 |  |  |  |  |  |  |  |  |  |  |
| +gp | 166 | 179 | 96 |  |  |  |  |  |  |  |  |  |  |
| TOTAL | 9823 | 11671 | 16085 |  |  |  |  |  |  |  |  |  |  |
| AGE\YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |
| 1 | 5544 | 2049 | 3545 | 9224 | 10191 | 7905 | 8204 | 12045 | 7287 | 3044 |  |  |  |
| 2 | 7099 | 4917 | 1818 | 3144 | 8181 | 9039 | 7011 | 7276 | 10683 | 6463 |  |  |  |
| 3 | 3079 | 4487 | 3587 | 1329 | 2082 | 6593 | 6641 | 5556 | 6045 | 8564 |  |  |  |
| 4 | 1154 | 1594 | 1992 | 2070 | 652 | 982 | 3490 | 3446 | 3155 | 3384 |  |  |  |
| 5 | 393 | 404 | 773 | 867 | 1057 | 255 | 501 | 1841 | 1350 | 1321 |  |  |  |
| 6 | 125 | 211 | 256 | 268 | 419 | 587 | 130 | 297 | 906 | 621 |  |  |  |
| 7 | 48 | 91 | 136 | 146 | 151 | 174 | 354 | 60 | 117 | 414 |  |  |  |
| 8 | 42 | 27 | 31 | 73 | 88 | 68 | 96 | 268 | 32 | 42 |  |  |  |
| +gp | 83 | 109 | 132 | 62 | 89 | 103 | 72 | 162 | 146 | 40 |  |  |  |
| TOTAL | 17568 | 13890 | 12269 | 17183 | 22910 | 25706 | 26498 | 30952 | 29721 | 23895 |  |  |  |
| AGE\YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |
| 1 | 2184 | 4778 | 4500 | 2896 | 3951 | 5242 | 3827 | 3338 | 2235 | 2199 |  |  |  |
| 2 | 2700 | 1937 | 4238 | 3991 | 2569 | 3504 | 4649 | 3395 | 2961 | 1982 |  |  |  |
| 3 | 4982 | 2240 | 1455 | 3005 | 2580 | 1875 | 2648 | 3358 | 2615 | 2225 |  |  |  |
| 4 | 4253 | 2461 | 977 | 795 | 1555 | 1407 | 1125 | 1657 | 1737 | 1452 |  |  |  |
| 5 | 1300 | 1486 | 1060 | 530 | 438 | 692 | 548 | 513 | 595 | 741 |  |  |  |
| 6 | 474 | 536 | 774 | 496 | 340 | 234 | 299 | 280 | 199 | 258 |  |  |  |
| 7 | 400 | 192 | 307 | 427 | 266 | 191 | 97 | 136 | 134 | 86 |  |  |  |
| 8 | 275 | 274 | 82 | 197 | 271 | 155 | 116 | 31 | 64 | 42 |  |  |  |
| +gp | 180 | 276 | 374 | 189 | 187 | 421 | 299 | 124 | 91 | 104 |  |  |  |
| TOTAL | 16747 | 14181 | 13767 | 12528 | 12156 | 13720 | 13608 | 12833 | 10633 | 9089 |  |  |  |
| AGE\YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | GMST 77-08 | AMST 77-08 |
| 1 | 3229 | 2280 | 2039 | 1261 | 2034 | 2814 | 2156 | 3535 | 2191 | 3021 | 0 | 3846 | 4519 |
| 2 | 1950 | 2864 | 2022 | 1809 | 1118 | 1804 | 2495 | 1912 | 3136 | 1943 | 2679 | 3425 | 4021 |
| 3 | 1529 | 1428 | 1927 | 1634 | 1379 | 873 | 1410 | 1901 | 1573 | 2527 | 1608 | 2503 | 2979 |
| 4 | 1049 | 786 | 918 | 1087 | 912 | 678 | 466 | 819 | 1203 | 1074 | 1682 | 1314 | 1564 |
| 5 | 533 | 476 | 393 | 302 | 527 | 406 | 352 | 281 | 406 | 650 | 639 | 568 | 669 |
| 6 | 307 | 281 | 236 | 177 | 127 | 221 | 234 | 186 | 152 | 224 | 329 | 276 | 320 |
| 7 | 119 | 136 | 148 | 103 | 76 | 70 | 128 | 136 | 93 | 90 | 135 | 138 | 164 |
| 8 | 34 | 56 | 62 | 70 | 59 | 43 | 39 | 66 | 77 | 56 | 60 | 70 | 93 |
| +gp | 50 | 69 | 71 | 64 | 64 | 51 | 48 | 32 | 104 | 66 | 88 |  |  |
| TOTAL | 8800 | 8377 | 7817 | 6506 | 6297 | 6959 | 7328 | 8870 | 8935 | 9651 | 7222 |  |  |

Table 7.10.13. Plaice in Divisions VIIf\&g. Summary.

Run title : CELTIC SEA PLAICE 2010 WG COMBSEX PLUSGROUP

At 15/05/2010 10:56

Table 16 Summary (without SOP correction)

Terminal Fs derived using XSA (With F shrinkage)

RECRUITS TOTALBIO TOTSPBIO LANDINGS YIELD/SSB FBAR 3-6 Age 1

| 1977 | 3582 | 2345 | 1169 | 757 | 0.6473 | 0.632 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 4964 | 2131 | 1010 | 875 | 0.8662 | 0.673 |
| 1979 | 8004 | 3238 | 1323 | 863 | 0.6524 | 0.6664 |
| 1980 | 5544 | 4078 | 1789 | 1373 | 0.7676 | 0.5416 |
| 1981 | 2049 | 3200 | 1792 | 1377 | 0.7683 | 0.4879 |
| 1982 | 3545 | 3255 | 2054 | 1303 | 0.6344 | 0.6304 |
| 1983 | 9224 | 3682 | 1938 | 1146 | 0.5912 | 0.5515 |
| 1984 | 10191 | 7100 | 2297 | 1210 | 0.5268 | 0.6686 |
| 1985 | 7905 | 5547 | 2635 | 1752 | 0.665 | 0.5021 |
| 1986 | 8204 | 4185 | 2814 | 1691 | 0.601 | 0.5285 |
| 1987 | 12045 | 6042 | 3245 | 1901 | 0.5858 | 0.6667 |
| 1988 | 7287 | 8812 | 3801 | 2116 | 0.5566 | 0.6323 |
| 1989 | 3044 | 5029 | 3140 | 2151 | 0.6851 | 0.6605 |
| 1990 | 2184 | 5272 | 3380 | 2082 | 0.616 | 0.7659 |
| 1991 | 4778 | 4339 | 2744 | 1501 | 0.547 | 0.6008 |
| 1992 | 4500 | 3803 | 2479 | 1188 | 0.4792 | 0.5219 |
| 1993 | 2896 | 3470 | 2008 | 1114 | 0.5548 | 0.4614 |
| 1994 | 3951 | 3113 | 1927 | 1070 | 0.5552 | 0.5348 |
| 1995 | 5242 | 3964 | 1956 | 1028 | 0.5257 | 0.6726 |
| 1996 | 3827 | 3812 | 1780 | 952 | 0.5349 | 0.5578 |
| 1997 | 3338 | 3340 | 1768 | 1217 | 0.6882 | 0.7211 |
| 1998 | 2235 | 2932 | 1659 | 1067 | 0.6433 | 0.6586 |
| 1999 | 2199 | 1994 | 1363 | 968 | 0.7104 | 0.7323 |
| 2000 | 3229 | 2020 | 1165 | 718 | 0.6161 | 0.6076 |
| 2001 | 2280 | 2596 | 1241 | 714 | 0.5753 | 0.5008 |
| 2002 | 2039 | 1762 | 1048 | 642 | 0.6123 | 0.7087 |
| 2003 | 1261 | 1783 | 1038 | 594 | 0.5724 | 0.6325 |
| 2004 | 2034 | 1426 | 863 | 510 | 0.5907 | 0.6269 |
| 2005 | 2814 | 1637 | 798 | 386 | 0.4836 | 0.476 |
| 2006 | 2156 | 2084 | 842 | 404 | 0.4798 | 0.4362 |
| 2007 | 3535 | 1878 | 859 | 410 | 0.4775 | 0.4969 |
| 2008 | 2191 | 2104 | 1063 | 437 | 0.4111 | 0.4084 |
| 2009 | 3021 | 2231 | 1128 | 463 | 0.4104 | 0.4076 |

Arith.

| Mean <br> O Units | 4403 <br> (Thousands) | 3461 <br> (Tonnes) | 1822 <br> (Tonnes) | 1090 <br> (Tonnes) | 0.5949 | 0.587 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |

Table 7.10.14. Plaice in Divisions VIIf\&g. Short-term forecast input data.
MFDP version 1a
Run: Plaice_Vllfg_sq
Time and date: 17:19 15/05/2010
Fbar age range: 3-6

| 2010 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  | SWt | Sel | cWt |
|  | 1 | 2815 | 0.12 | 0 | 0 | 0 | 0.15 | 0 | 0.171 |
|  | 2 | 2679 | 0.12 | 0.26 | 0 | 0 | 0.195 | 0.08 | 0.220667 |
|  | 3 | 1608 | 0.12 | 0.52 | 0 | 0 | 0.248 | 0.295467 | 0.277667 |
|  | 4 | 1682 | 0.12 | 0.86 | 0 | 0 | 0.300667 | 0.492233 | 0.319667 |
|  | 5 | 639 | 0.12 | 1 | 0 | 0 | 0.343333 | 0.511533 | 0.371667 |
|  | 6 | 329 | 0.12 | 1 | 0 | 0 | 0.405333 | 0.4513 | 0.444 |
|  | 7 | 135 | 0.12 | 1 | 0 | 0 | 0.487333 | 0.3728 | 0.535333 |
|  | 8 | 60 | 0.12 | 1 | 0 | 0 | 0.588 | 0.306367 | 0.646 |
|  | 9 | 88 | 0.12 | 1 | 0 | 0 | 0.810946 | 0.306367 | 0.88495 |


| 2011 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  | Mat |  |  | SWt | Sel | Wt |
|  | 1 | 2815 | 0.12 | 0 | 0 | 0 | 0.15 | 0 | 0.171 |
|  | 2 |  | 0.12 | 0.26 | 0 | 0 | 0.195 | 0.08 | 0.220667 |
|  | 3 |  | 0.12 | 0.52 | 0 | 0 | 0.248 | 0.295467 | 0.277667 |
|  | 4 |  | 0.12 | 0.86 | 0 | 0 | 0.300667 | 0.492233 | 0.319667 |
|  | 5 |  | 0.12 | 1 | 0 | 0 | 0.343333 | 0.511533 | 0.371667 |
|  | 6 |  | 0.12 | 1 | 0 | 0 | 0.405333 | 0.4513 | 0.444 |
|  | 7 |  | 0.12 | 1 | 0 | 0 | 0.487333 | 0.3728 | 0.535333 |
|  | 8 |  | 0.12 | 1 | 0 | 0 | 0.588 | 0.306367 | 0.646 |
|  | 9 |  | 0.12 | 1 | 0 | 0 | 0.810946 | 0.306367 | 0.88495 |



Input units are thousands and kg - output in tonnes

Table 7.10.15. Plaice in Divisions VIIf\&g. Management option table status quo forecast.
MFDP version 1a
Run: Plaice_Vllfg_sq
CELTIC SEA PLAICE
Time and date: 17:19 15/05/2010
Fbar age range: 3-6

| 2010 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 2374 | 1303 | 1.0000 | 0.4376 | 539 |


| 2011 <br> Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2440 | 1352 | 0.0000 | 0.0000 | 0 | 3067 | 1939 |
| . | 1352 | 0.1000 | 0.0438 | 66 | 3000 | 1876 |
| . | 1352 | 0.2000 | 0.0875 | 129 | 2935 | 1816 |
| . | 1352 | 0.3000 | 0.1313 | 190 | 2873 | 1758 |
| . | 1352 | 0.4000 | 0.1751 | 249 | 2813 | 1703 |
| . | 1352 | 0.5000 | 0.2188 | 305 | 2756 | 1650 |
| . | 1352 | 0.6000 | 0.2626 | 359 | 2701 | 1598 |
| . | 1352 | 0.7000 | 0.3063 | 411 | 2648 | 1549 |
| . | 1352 | 0.8000 | 0.3501 | 461 | 2597 | 1502 |
| . | 1352 | 0.9000 | 0.3939 | 510 | 2548 | 1457 |
| . | 1352 | 1.0000 | 0.4376 | 556 | 2501 | 1414 |
| . | 1352 | 1.1000 | 0.4814 | 601 | 2455 | 1372 |
| . | 1352 | 1.2000 | 0.5252 | 644 | 2411 | 1332 |
| . | 1352 | 1.3000 | 0.5689 | 686 | 2369 | 1293 |
| . | 1352 | 1.4000 | 0.6127 | 726 | 2329 | 1256 |
| . | 1352 | 1.5000 | 0.6565 | 765 | 2290 | 1221 |
| . | 1352 | 1.6000 | 0.7002 | 802 | 2252 | 1187 |
| . | 1352 | 1.7000 | 0.7440 | 838 | 2216 | 1154 |
| . | 1352 | 1.8000 | 0.7877 | 873 | 2181 | 1122 |
| . | 1352 | 1.9000 | 0.8315 | 906 | 2147 | 1092 |
| . | 1352 | 2.0000 | 0.8753 | 938 | 2115 | 1062 |

Input units are thousands and kg - output in tonnes

Table 7.10.16. Plaice in Divisions VIIf\&g. Forecast detailed results; status quo forecast.

MFDP version 1a
Run: Plaice_Vllfg_sq
Time and date: 17:19 15/05/2010
Fbar age range: 3-6

| Year: Age |  | 2010 | F multiplier | 1 Fbar: |  | 0.4376 | SSNos(Jar SSB(Jan) |  | SSNos(ST' SSB(ST) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CatchNos Yield |  | StockNos | Biomass |  |  |  |  |
|  | 1 | 0 | 0 | 0 | 02815 | 422 | 0 | 0 | 0 | 0 |
|  | 2 | 0.08 | 194 | 43 | 3679 | 522 | 697 | 136 | 697 | 136 |
|  | 3 | 0.2955 | 389 | 108 | 1608 | 399 | 836 | 207 | 836 | 207 |
|  | 4 | 0.4922 | 619 | 198 | 1682 | 506 | 1447 | 435 | 1447 | 435 |
|  | 5 | 0.5115 | 242 | 90 | 036 | 219 | 639 | 219 | 639 | 219 |
|  | 6 | 0.4513 | 113 | 50 | 329 | 133 | 329 | 133 | 329 | 133 |
|  | 7 | 0.3728 | 40 | 21 | 1135 | 66 | 135 | 66 | 135 | 66 |
|  | 8 | 0.3064 | 15 | 10 | 060 | 35 | 60 | 35 | 60 | 35 |
|  | 9 | 0.3064 | 22 | 19 | -88 | 71 | 88 | 71 | 88 | 71 |
| Total |  |  | 1634 | 539 | 10035 | 2374 | 4230 | 1303 | 4230 | 1303 |
| Year: |  | 2011 | F multiplier |  | 1 Fbar: | 0.4376 |  |  |  |  |
| Age |  |  | CatchNos Yield |  | StockNos | Biomass | SSNos(Jar | SSB(Jan) | SSNos(ST) | SSB(ST) |
|  | 1 | 0 | 0 | 0 | 02815 | 422 | 0 | 0 | 0 | 0 |
|  | 2 | 0.08 | 181 | 40 | 2497 | 487 | 649 | 127 | 649 | 127 |
|  | 3 | 0.2955 | 530 | 147 | 2193 | 544 | 1141 | 283 | 1141 | 283 |
|  | 4 | 0.4922 | 391 | 125 | 51061 | 319 | 913 | 274 | 913 | 274 |
|  | 5 | 0.5115 | 346 | 129 | 912 | 313 | 912 | 313 | 912 | 313 |
|  | 6 | 0.4513 | 117 | 52 | 340 | 138 | 340 | 138 | 340 | 138 |
|  | 7 | 0.3728 | 55 | 29 | 186 | 91 | 186 | 91 | 186 | 91 |
|  | 8 | 0.3064 | 21 | 13 | 382 | 48 | 82 | 48 | 82 | 48 |
|  | 9 | 0.3064 | 24 | 21 | 197 | 78 | 97 | 78 | 97 | 78 |
| Total |  |  | 1664 | 556 | 610183 | 2440 | 4319 | 1352 | 4319 | 1352 |
| Year: |  | 2012 | F multiplier |  | 1 Fbar: | 0.4376 |  |  |  |  |
| Age |  |  | CatchNos Yield |  | StockNos | Biomass | SSNos(Jar | SSB(Jan) | SSNos(ST) | SSB(ST) |
|  | 1 | 0 | 0 | 0 | 02815 | 422 | 0 | 0 | 0 | 0 |
|  | 2 | 0.08 | 181 | 40 | 2497 | 487 | 649 | 127 | 649 | 127 |
|  | 3 | 0.2955 | 494 | 137 | 2044 | 507 | 1063 | 264 | 1063 | 264 |
|  | 4 | 0.4922 | 533 | 170 | 1448 | 435 | 1245 | 374 | 1245 | 374 |
|  | 5 | 0.5115 | 218 | 81 | 81575 | 198 | 575 | 198 | 575 | 198 |
|  | 6 | 0.4513 | 167 | 74 | 4485 | 197 | 485 | 197 | 485 | 197 |
|  | 7 | 0.3728 | 56 | 30 | 192 | 94 | 192 | 94 | 192 | 94 |
|  | 8 | 0.3064 | 28 | 18 | 18114 | 67 | 114 | 67 | 114 | 67 |
|  | 9 | 0.3064 | 29 | 26 | 6117 | 95 | 117 | 95 | 117 | 95 |
| Total |  |  | 1707 | 577 | 10286 | 2501 | 4440 | 1414 | 4440 | 1414 |

Input units are thousands and kg - output in tonnes

Table 7.1.17. Plaice in Divisions VIIf\&g. Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to ladings and SSB (by weight) of these year classes.

| Y ear-class |  |  | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock No. (thousands) |  |  | 3535 | 2191 | 3021 | 2815 | 2815 |
| of |  | year-olds |  |  |  |  |  |
| Source |  |  | XSA | XSA | XSA | GM89-08 | GM89-08 |
| Status Quo F: |  |  |  |  |  |  |  |
| \% in | 2010 | landings | 36.7 | 20.0 | 8.0 | 0.0 | - |
| \% in | 2011 | landings | 23.2 | 22.5 | 26.4 | 7.2 | 0.0 |
| \% in | 2010 | SSB | 33.4 | 15.9 | 10.4 | 0.0 | - |
| \% in | 2011 | SSB | 23.2 | 20.3 | 20.9 | 9.4 | 0.0 |
| \% in | 2012 | SSB | 13.9 | 14.0 | 26.4 | 18.6 | 9.0 |

Plaice in VIlfg : Year-class \% contribution to


Table 7.10.18. Plaice in Divisions VIIf\&g. Yield-per-recruit summary table.
MFYPR version $2 a$
Run: Plaice_VIlfg_yield
Time and date: 17:33 15/05/2010
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 8.8433 | 4.3997 | 6.7118 | 3.9987 | 6.7118 | 3.9987 |
| 0.1000 | 0.0438 | 0.1895 | 0.1034 | 7.2667 | 3.2359 | 5.1418 | 2.8367 | 5.1418 | 2.8367 |
| 0.2000 | 0.0875 | 0.3108 | 0.1573 | 6.2584 | 2.5186 | 4.1399 | 2.1212 | 4.1399 | 2.1212 |
| 0.3000 | 0.1313 | 0.3942 | 0.1865 | 5.5664 | 2.0455 | 3.4542 | 1.6499 | 3.4542 | 1.6499 |
| 0.4000 | 0.1751 | 0.4544 | 0.2024 | 5.0670 | 1.7182 | 2.9609 | 1.3243 | 2.9609 | 1.3243 |
| 0.5000 | 0.2188 | 0.4997 | 0.2110 | 4.6924 | 1.4832 | 2.5923 | 1.0909 | 2.5923 | 1.0909 |
| 0.6000 | 0.2626 | 0.5347 | 0.2153 | 4.4028 | 1.3095 | 2.3086 | 0.9188 | 2.3086 | 0.9188 |
| 0.7000 | 0.3063 | 0.5626 | 0.2172 | 4.1732 | 1.1778 | 2.0847 | 0.7887 | 2.0847 | 0.7887 |
| 0.8000 | 0.3501 | 0.5852 | 0.2178 | 3.9873 | 1.0759 | 1.9045 | 0.6884 | 1.9045 | 0.6884 |
| 0.9000 | 0.3939 | 0.6039 | 0.2175 | 3.8341 | 0.9956 | 1.7567 | 0.6095 | 1.7567 | 0.6095 |
| 1.0000 | 0.4376 | 0.6196 | 0.2168 | 3.7057 | 0.9311 | 1.6338 | 0.5465 | 1.6338 | 0.5465 |
| 1.1000 | 0.4814 | 0.6329 | 0.2159 | 3.5968 | 0.8787 | 1.5301 | 0.4955 | 1.5301 | 0.4955 |
| 1.2000 | 0.5252 | 0.6444 | 0.2150 | 3.5031 | 0.8354 | 1.4415 | 0.4536 | 1.4415 | 0.4536 |
| 1.3000 | 0.5689 | 0.6545 | 0.2140 | 3.4216 | 0.7991 | 1.3651 | 0.4187 | 1.3651 | 0.4187 |
| 1.4000 | 0.6127 | 0.6633 | 0.2131 | 3.3501 | 0.7685 | 1.2985 | 0.3894 | 1.2985 | 0.3894 |
| 1.5000 | 0.6565 | 0.6712 | 0.2122 | 3.2868 | 0.7422 | 1.2400 | 0.3645 | 1.2400 | 0.3645 |
| 1.6000 | 0.7002 | 0.6782 | 0.2114 | 3.2303 | 0.7195 | 1.1882 | 0.3431 | 1.1882 | 0.3431 |
| 1.7000 | 0.7440 | 0.6845 | 0.2107 | 3.1794 | 0.6997 | 1.1419 | 0.3245 | 1.1419 | 0.3245 |
| 1.8000 | 0.7877 | 0.6903 | 0.2101 | 3.1334 | 0.6823 | 1.1004 | 0.3083 | 1.1004 | 0.3083 |
| 1.9000 | 0.8315 | 0.6955 | 0.2094 | 3.0914 | 0.6668 | 1.0629 | 0.2940 | 1.0629 | 0.2940 |
| 2.0000 | 0.8753 | 0.7003 | 0.2089 | 3.0530 | 0.6529 | 1.0289 | 0.2813 | 1.0289 | 0.2813 |


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.4376 |
| FMax | 0.8074 | 0.3534 |
| F0.1 | 0.3654 | 0.1599 |
| F35\%SPR | 0.3736 | 0.1635 |

Table 7.10.19. Plaice in Divisions VIIf\&g. MSY analysis sen file.


Table 7.10.20. Plaice in Divisions VIIf\&g. MSY analysis sum file.

| $12$ |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1977 | 2009 |  |  |  |  |  |  |  |  |  |  |
| Recruits, age 1, (thousands) |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1000 |  |  |  |  |  |  |  |  |  |  |
| SSB, (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| TSB, (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Catch, Total (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Catch, H.cons (tonnes) |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |
| Mean F, Total |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 6 |  |  |  |  |  |  |  |  |  |  |
| Mean F, H.cons. |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 6 |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| Not used |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 1977 | 3582 | 1169 | 2345 | 757 | 757 | 0 | 0 | 0.632 | 0.632 | 0 | 0 |
| 1978 | 4964 | 1010 | 2131 | 875 | 875 | 0 | 0 | 0.673 | 0.673 | 0 | 0 |
| 1979 | 8004 | 1323 | 3238 | 863 | 863 | 0 | 0 | 0.6664 | 0.6664 | 0 | 0 |
| 1980 | 5544 | 1789 | 4078 | 1373 | 1373 | 0 | 0 | 0.5416 | 0.5416 | 0 | 0 |
| 1981 | 2049 | 1792 | 3200 | 1377 | 1377 | 0 | 0 | 0.4879 | 0.4879 | 0 | 0 |
| 1982 | 3545 | 2054 | 3255 | 1303 | 1303 | 0 | 0 | 0.6304 | 0.6304 | 0 | 0 |
| 1983 | 9224 | 1938 | 3682 | 1146 | 1146 | 0 | 0 | 0.5515 | 0.5515 | 0 | 0 |
| 1984 | 10191 | 2297 | 7100 | 1210 | 1210 | 0 | 0 | 0.6686 | 0.6686 | 0 | 0 |
| 1985 | 7905 | 2635 | 5547 | 1752 | 1752 | 0 | 0 | 0.5021 | 0.5021 | 0 | 0 |
| 1986 | 8204 | 2814 | 4185 | 1691 | 1691 | 0 | 0 | 0.5285 | 0.5285 | 0 | 0 |
| 1987 | 12045 | 3245 | 6042 | 1901 | 1901 | 0 | 0 | 0.6667 | 0.6667 | 0 | 0 |
| 1988 | 7287 | 3801 | 8812 | 2116 | 2116 | 0 | 0 | 0.6323 | 0.6323 | 0 | 0 |
| 1989 | 3044 | 3140 | 5029 | 2151 | 2151 | 0 | 0 | 0.6605 | 0.6605 | 0 | 0 |
| 1990 | 2184 | 3380 | 5272 | 2082 | 2082 | 0 | 0 | 0.7659 | 0.7659 | 0 | 0 |
| 1991 | 4778 | 2744 | 4339 | 1501 | 1501 | 0 | 0 | 0.6008 | 0.6008 | 0 | 0 |
| 1992 | 4500 | 2479 | 3803 | 1188 | 1188 | 0 | 0 | 0.5219 | 0.5219 | 0 | 0 |
| 1993 | 2896 | 2008 | 3470 | 1114 | 1114 | 0 | 0 | 0.4614 | 0.4614 | 0 | 0 |
| 1994 | 3951 | 1927 | 3113 | 1070 | 1070 | 0 | 0 | 0.5348 | 0.5348 | 0 | 0 |
| 1995 | 5242 | 1956 | 3964 | 1028 | 1028 | 0 | 0 | 0.6726 | 0.6726 | 0 | 0 |
| 1996 | 3827 | 1780 | 3812 | 952 | 952 | 0 | 0 | 0.5578 | 0.5578 | 0 | 0 |
| 1997 | 3338 | 1768 | 3340 | 1217 | 1217 | 0 | 0 | 0.7211 | 0.7211 | 0 | 0 |
| 1998 | 2235 | 1659 | 2932 | 1067 | 1067 | 0 | 0 | 0.6586 | 0.6586 | 0 | 0 |
| 1999 | 2199 | 1363 | 1994 | 968 | 968 | 0 | 0 | 0.7323 | 0.7323 | 0 | 0 |
| 2000 | 3229 | 1165 | 2020 | 718 | 718 | 0 | 0 | 0.6076 | 0.6076 | 0 | 0 |
| 2001 | 2280 | 1241 | 2596 | 714 | 714 | 0 | 0 | 0.5008 | 0.5008 | 0 | 0 |
| 2002 | 2039 | 1048 | 1762 | 642 | 642 | 0 | 0 | 0.7087 | 0.7087 | 0 | 0 |
| 2003 | 1261 | 1038 | 1783 | 594 | 594 | 0 | 0 | 0.6325 | 0.6325 | 0 | 0 |
| 2004 | 2034 | 863 | 1426 | 510 | 510 | 0 | 0 | 0.6269 | 0.6269 | 0 | 0 |
| 2005 | 2814 | 798 | 1637 | 386 | 386 | 0 | 0 | 0.476 | 0.476 | 0 | 0 |
| 2006 | 2156 | 842 | 2084 | 404 | 404 | 0 | 0 | 0.4362 | 0.4362 | 0 | 0 |
| 2007 | 3535 | 859 | 1878 | 410 | 410 | 0 | 0 | 0.4969 | 0.4969 | 0 | 0 |
| 2008 | 2191 | 1063 | 2104 | 437 | 437 | 0 | 0 | 0.4084 | 0.4084 | 0 | 0 |
| 2009 | 3021 | 1128 | 2231 | 463 | 463 | 0 | 0 | 0.4075 | 0.4075 | 0 | 0 |

Table 7.10.21. Plaice in Divisions VIIfg. Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships and the yield-per-recruit $F_{\text {msy }}$ proxies.

```
Stock name
Plaice VIIfg
Sen filename
wgcse_ple-celt.sen
pf, pm
    0 0
Number of iterations
        1000
Simulate variation in Biological parameters
    TRUE
SR relationship constrained
    TRUE
```

Ricker
700/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Determinis | 0.98 | 0.35 | 3631 | 1164 | 1.12 | 0.32 | 3.75 | 0.00028 | 54.04 |
| Mean | 1.23 | 0.38 | 4515 | 1387 | 1.15 | 0.34 | 3.96 | 0.00029 |  |
| 5\%ile | 0.57 | 0.25 | 2186 | 818 | 0.95 | 0.13 | 2.66 | 0.00011 |  |
| 25\%ile | 0.78 | 0.31 | 2801 | 1017 | 1.05 | 0.26 | 3.34 | 0.00022 |  |
| 50\%ile | 1.01 | 0.36 | 3599 | 1180 | 1.13 | 0.33 | 3.84 | 0.00029 |  |
| 75\%ile | 1.45 | 0.43 | 4692 | 1462 | 1.23 | 0.43 | 4.51 | 0.00037 |  |
| 95\%ile | 2.73 | 0.56 | 8981 | 2337 | 1.38 | 0.55 | 5.50 | 0.00047 |  |
| CV | 0.59 | 0.25 | 1.00 | 0.73 | 0.12 | 0.38 | 0.22 | 0.38 |  |

Beverton-Holt
616/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Determinis | 1.20 | 0.19 | 8666 | 1484 | 0.35 | 0.90 | 8097 | 1841 | 54.38 |
| Mean | 1.36 | 0.18 | 37227 | 2313 | 0.33 | 0.94 | 11735 | 3550 |  |
| 5\%ile | 0.53 | 0.03 | 4001 | 876 | 0.10 | 0.78 | 4955 | 594 |  |
| 25\%ile | 0.79 | 0.14 | 7034 | 1234 | 0.23 | 0.86 | 6672 | 1304 |  |
| 50\%ile | 1.11 | 0.19 | 10303 | 1653 | 0.33 | 0.93 | 8585 | 2083 |  |
| 75\%ile | 1.64 | 0.22 | 23139 | 2430 | 0.42 | 1.00 | 12108 | 3768 |  |
| 95\%ile | 3.12 | 0.28 | 171972 | 5586 | 0.57 | 1.13 | 27540 | 10160 |  |
| CV | 0.60 | 0.43 | 2.15 | 1.02 | 0.42 | 0.11 | 0.98 | 1.51 |  |

Smooth hockeystick
719/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Determinis | 0.62 | 0.35 | 3790 | 1193 | 0.51 | 1.13 | 1.24 | 2059 | 54.55 |
| Mean | 0.72 | 0.34 | 14968 | 1352 | 0.55 | 1.14 | 1.32 | 2069 |  |
| 5\%ile | 0.43 | 0.03 | 1889 | 834 | 0.44 | 0.67 | 1.05 | 1221 |  |
| 25\%ile | 0.56 | 0.23 | 2722 | 1090 | 0.49 | 0.90 | 1.18 | 1640 |  |
| 50\%ile | 0.66 | 0.34 | 4189 | 1270 | 0.53 | 1.13 | 1.28 | 2062 |  |
| 75\%ile | 0.80 | 0.46 | 7107 | 1537 | 0.59 | 1.31 | 1.42 | 2395 |  |
| 95\%ile | 1.19 | 0.66 | 85576 | 2200 | 0.72 | 1.67 | 1.74 | 3047 |  |
| CV | 0.38 | 0.57 | 2.08 | 0.29 | 0.17 | 0.27 | 0.17 | 0.27 |  |





Figure 7.10.1. Plaice in Division VIIf\&g. UK (E\&W) lpue and effort by fleet.


Figure 7.10.2. Plaice in Division VIIf\&g. Ireland and Belgium: lpue and effort by fleet.


Figure 7.10.3a. Plaice in Division VIIf\&g. Ireland otter trawl discard sampling results in 2007-2009: raised to sampled trips.


Figure 7.10.3b. Plaice in Division VIIf\&g. UK (E\&W) Discard sampling results in 2009: raised to sampled trips. All gears


Figure 7.10.4. Plaice in Division VIIf\&g. Length distributions of UK (England \& Wales) landings from 2000 to 2009.


Figure 7.10.4. Plaice in Division VIIf\&g. Age composition of International landings from 2000 to 2009.


Figure 7.10.6. UK (BTS-Q3) Beam trawl survey log cpue by year, year class, log catch curves and the negative slope of the catch curves ( $\sim \mathrm{Z}$ ).


Figure 7.10.7a. UK EW Beam trawl fleet $\log$ cpue by year, year class, log catch curves and the negative slope of the catch curves ( $\sim \mathrm{Z}$ ).


Figure 7.10.7b. UK EW Otter trawl fleet $\log$ cpue by year, year class, log catch curves and the negative slope of the catch curves ( $\sim \mathrm{Z}$ ).


Figure 7.10.8. Plaice in Division VIIf\&g. Commercial fleet and survey $\log$ catchability residuals from the final run.


Figure 7.10.9. Plaice in Division VIIf\&g. Assessment model estimate retrospective bias.


Figure 7.10.10. Plaice in Division VIIf\&g. The time-series of stock and fishery trends.

Plaice in VIIfg. Stock-Recruitment



MFYPR version 2a
Run: Plaice_VIlfg_yield
Time and date: 17:33 15/05/2010

| Reference point | F multiplier | Absolute $\mathbf{F}$ |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.4376 |
| FMax | 0.8074 | 0.3534 |
| F0.1 | 0.3654 | 0.1599 |
| F35\%SPR | 0.3736 | 0.1635 |

MFDP version 1a
Run: Plaice_Vllfg_sq
CELTIC SEA PLAICE
Time and date: 17:19 15/05/2010
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Figure 7.10.12. Plaice in Division VIIf\&g. Yield-per-recruit and short-term forecast.


Figure 7.10.13. Plaice in Divisions VIIfg. MSY fitted stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $\mathrm{F}_{\text {msy }}$. Right hand panels: curves plotted from the first 100 MCMC resamples with converged $\mathrm{F}_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of Fmš from the $\mathbf{1 0 0 0}$ re-samples.

Plaice VIlfg Ricker


Figure 7.10.14. Plaice in Divisions VIIfg. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Ricker stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathbf{F}_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

## Plaice VIlfg Beverton-Holt



Figure 7.10.15. Plaice in Divisions VIIfg. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Beverton and Holt stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $F_{\text {msy. }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.


Figure 7.10.16. Plaice in Divisions VIIfg. Estimates of $F$ reference points and equilibrium yield and SSB against fishing mortality using smooth hockey stick stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $\mathrm{F}_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $\mathrm{F}_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

## Plaice VIlfg - Per recruit statistics



Figure 7.10.17. Plaice in Divisions VIIfg. Fitted yield-per-recruit F reference points, yield-perrecruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first 100 re-samples.

### 7.11 Plaice in the Southwest of Ireland (ICES Divisions VIIh-k)

## Type of assessment in 2010

No assessment was performed, however catch numbers and weights were aggregated for the Irish landings for the years 1993-2009 and these were used to perform a yield-per-recruit analysis for the VIIjk part of the stock.

### 7.11.1 General

## Stock Identity

Plaice in VIIj are mainly caught by Irish vessels on sandy grounds off counties Kerry and west Cork. Plaice catches in VIIk are negligible. VIIh is also considered part of the stock for assessment purposes but there is no evidence to suggest that this is actually the same stock.

### 7.11.2 Data

The nominal landings are given in Table 7.11.1.
Most non-Irish landings were from VIIh which is likely to be a different stock. Because age data were only available for Irish landings (which were mainly from VIIjk) the remainder of Section 7.11 concerns Irish data only in VIIjk

## Sampling

Figure 7.11 .1 shows that plaice landings in VIIjk in 2009 were mostly taken in VIIj by otter trawlers. This was reflected in the sampling.

## Data quality

Figure 7.11.2 shows the length distribution of the Irish landings in VIIjk between 1993 and 2009. Sample numbers appear to be adequate. There are no distinct modes of strong year classes discernible. One sample was removed (420-DEM196); it contained 192 plaice at 27 cm and no other length classes. In 1994 and 1995 a considerable number of small plaice $(<20 \mathrm{~cm})$ appeared in the samples. The most likely explanation for this is that discard fish were mistakenly entered as landings; these were therefore excluded from the analysis. The age data for 1995 were considered insufficient and for this year the combined age data for 1993-1996 were used.

Annual Age-Length-Keys (ALKs) were constructed (all quarters and gear types combined) and applied to the sampled length frequency distributions. Figure 7.11.3 shows the age distribution of plaice in VIIjk between 1993 and 2009.

### 7.11.3 Historical stock development

Because plaice in VIIh were not sampled, it would not be appropriate to raise the data to all landings in VIIhjk. Instead, the official International landings figures for VIIjk were used to raise the age distributions (Table 7.11.2).

The estimated catch numbers-at-age are given in Table 7.11.3, catch weights-at-age are given in Table 7.11.4. There appears to be relatively little contrast (particularly weak or strong year classes) in the catch numbers. This is also illustrated by Figure 7.11.4, which shows the standardised catch proportions-at-age. Figure 7.11.5 shows the log catch numbers-at-age. The rate of decline in catch numbers through the cohorts appears to be reasonably stable. This can be further investigated by calculating
the slope of the $\log$ catch numbers $(Z)$. Figure 7.11 .6 shows the catch curve; plaice under the age of 4 are not fully selected and from age 7 onwards the data get quite noisy, therefore the slope of the log catch numbers was estimated over ages 4 to 7 (Figure 7.11.7). It appears that Z varied between 0.6 and 1.2. The estimate for Z appears to be quite variable. These levels of Z are quite high compared to other plaice stocks. There is a possibility that this can be the consequence of migration. If young fish are highly concentrated in inshore areas and have a higher catchability than older fish, which might be distributed more widely further offshore, this could result in apparent high levels of $Z$. This possibility will be investigated further intersessionally.

## Yield-per-recruit

The yield-per-recruit was estimated using a method by Thompson and Bell (1934). This method requires the selectivity to be estimated. This was done by estimating the slope of the log catch numbers for ages that are fully selected and using this slope ( $Z$ ) to predict the population numbers for ages that are not fully selected. The Z was estimated on pseudo-cohorts which were standardised to take account of annual variations in the catch numbers. Figure 7.11 .8 shows that plaice in VIIjk appear to be fully selected by the age of 4 and that after the age of 9 the data get very sparse. Figure 7.11.9 shows the slope of the mean log standardised catch numbers. The predicted catch numbers from this slope were used to estimate the 'observed' selectivity. This was then modelled by applying a linear model after a logit transformation. The estimated selection curve is also shown in Figure 7.11.9. A natural mortality of 0.12 was assumed (based on the value used by the WG for plaice in VIIfg) and the WG maturity ogive for plaice in VIIfg was used to estimate SSB. The yield was estimated for a range of F values based on the average catch weights. Figure 7.11 .10 shows the YPR curve, $\mathrm{F}_{\max }$ is estimated to be $0.24 \mathrm{~F}_{0.1}$ is estimated at 0.14 . Recent values of Z ranged from 0.5 to 1.2, with $\mathrm{M}=0.12$ this would result in an F of between 0.48 and 1.08. This is well above $\mathrm{F}_{\max }$ and $\mathrm{F}_{0.1}$.

### 7.11.4 References

Thompson and Bell. 1934. W.F. Thompson and F.H. Bell, Biological statistics of the Pacific halibut fishery. 2. Effect of changes in intensity upon total yield and yield per unit of gear, Rep. Int. Fish. (Pacific Halibut) Comm. 8 (1934), p. 49.

Table 7.11.1. Plaice in Divisions VII h-k (Southwest Ireland). Nominal landings ( $\mathbf{t}$ ), 1987-2009, as officially reported to ICES.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium* | 250 | 245 | 403 | 301 | 252 | 246 | 344 | 197 | 235 |
| Denmark | 1 | 1 | 1 | - | - | - | - | - | - |
| France | 85 | 135 | 229 | 77 | 173 | 90 | 64 | 48 | 60 |
| Ireland | 300 | 369 | 454 | 338 | 478 | 477 | 383 | 271 | 321 |
| Netherlands | - | - | - | - | - | - | - | - | - |
| Spain | - | - | - | - | - | - | - | - | - |
| UK - Eng+Wales+N | . | . | 73 | 88 | 287 | 264 | 218 | 258 | 282 |
| UK - England \& Wa | 246 | 433 | . | . | . | . | . | . | . |
| UK - Scotland | - | 1 | - | 1 | 1 | 6 | 7 | 1 | 4 |
| Total | 882 | 1184 | 1160 | 805 | 1191 | 1083 | 1016 | 775 | 902 |
| Unallocated |  |  |  |  |  |  | -361 | -198 | -360 |
| WG estimate |  |  |  |  |  |  | 655 | 577 | 542 |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| Belgium* | 304 | 442 | 335 | 45 | 4 | 27 | 69 | 20 | 67 |
| Denmark | - | - | - | - | - | - | - | - | - |
| France | 48 | 69 | 49 | . | 54 | 50 | 45 | 32 | 32 |
| Ireland | 305 | 344 | 286 | 299 | 200 | 160 | 155 | 127 | 91 |
| Netherlands | 52 | - | 13 | 1 | 2 | - | - | - | - |
| Spain | - | - | - | 1 | 5 | 3 | 2 | 6 | 6 |
| UK - Eng+Wales+N | 154 | 138 | 106 | 82 | 75 | 73 | 59 | 56 | 36 |
| UK - England \& Wa | . | . | . |  | . | . | . | . | . |
| UK - Scotland | 1 | 1 | 1 | 1 | 1 | - | - | - | - |
| Total | 864 | 994 | 790 | 428 | 341 | 313 | 330 | 241 | 232 |
| Unallocated | -411 | -349 | -346 | -22 | -42 | -52 | -17 | -24 | -11 |
| WG estimate | 453 | 645 | 444 | 406 | 299 | 261 | 313 | 217 | 221 |
| Country | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |  |  |
| Belgium | 32 | 22 | 7 | 25 | 1 |  |  |  |  |
| Denmark |  |  |  |  |  |  |  |  |  |
| France | 20 | 37 | 30 | 12 |  |  |  |  |  |
| Ireland | 90 | 65 | 72 | 72 | 72 |  |  |  |  |
| Netherlands | . |  |  |  |  |  |  |  |  |
| Spain |  | 1 | 13 | 1 |  |  |  |  |  |
| UK - Eng+Wales+N | 28 | 18 | 20 | 12 | 32 |  |  |  |  |
| UK - England \& Wa | . |  |  |  |  |  |  |  |  |
| UK - Scotland | . |  |  |  |  |  |  |  |  |
| Total | 170 | 143 | 142 | 122 | 105 |  |  |  |  |
| Unallocated | -6 | 4 | -22 | 13 |  |  |  |  |  |
| WG estimate | 164 | 147 | 120 | 135 |  |  |  |  |  |

* Belgian Landings up to 1998 include VIIg

Table 7.11.2. Official landings $(\mathbf{t})$ of plaice in VIIjk.

| Year | Bel | Fra | IrI | Esp | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | . | 8 | 383 | - | 46 | 437 |
| 1994 | . | 6 | 251 | - | 60 | 317 |
| 1995 | . | 12 | 317 | - | 90 | 419 |
| 1996 | . | 3 | 295 | - | 38 | 336 |
| 1997 | . | 6 | 337 | - | 32 | 375 |
| 1998 | . | 8 | 282 | - | 16 | 306 |
| 1999 | 42 | 0 | 296 | $<0.5$ | 15 | 353 |
| 2000 | 4 | 16 | 195 | 5 | 9 | 229 |
| 2001 | - | 16 | 157 | 3 | 6 | 182 |
| 2002 | 14 | 21 | 155 | 2 | 5 | 197 |
| 2003 | 4 | 7 | 125 | 6 | 9 | 151 |
| 2004 | <0.5 | 5 | 87 | 6 | 6 | 104 |
| 2005 | - | 4 | 88 | - | 2 | 94 |
| 2006 | - | 6 | 63 | 1 | 1 | 71 |
| 2007 | - | 9 | 72 | 11 | 2 | 94 |
| 2008 | - | 5 | 72 | 1 | 1 | 79 |
| 2009* | - |  | 72 |  | 2 | 74 |

* Preliminary data

Table 7.11.3. Catch numbers-at-age for plaice in VIIjk.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 92 | 624 | 479 | 115 | 45 | 23 | 10 | 6 | 2 | 0 | 1 |
| 1994 | 68 | 104 | 340 | 260 | 82 | 46 | 18 | 8 | 5 | 1 | 1 | 0 |
| 1995 | 10 | 208 | 634 | 348 | 107 | 36 | 16 | 7 | 4 | 1 | 2 | 0 |
| 1996 | 1 | 77 | 316 | 229 | 127 | 37 | 23 | 5 | 1 | 0 | 0 | 0 |
| 1997 | 0 | 164 | 277 | 269 | 120 | 42 | 20 | 5 | 0 | 0 | 0 | 9 |
| 1998 | 0 | 46 | 355 | 164 | 103 | 38 | 26 | 10 | 4 | 3 | 0 | 0 |
| 1999 | 11 | 143 | 312 | 201 | 65 | 37 | 18 | 11 | 9 | 2 | 2 | 8 |
| 2000 | 2 | 74 | 161 | 190 | 64 | 36 | 7 | 5 | 3 | 2 | 0 | 2 |
| 2001 | 1 | 55 | 165 | 146 | 47 | 6 | 21 | 2 | 7 | 0 | 0 | 0 |
| 2002 | 0 | 54 | 155 | 172 | 54 | 42 | 44 | 12 | 4 | 2 | 0 | 1 |
| 2003 | 0 | 74 | 165 | 65 | 29 | 6 | 15 | 11 | 2 | 2 | 1 | 0 |
| 2004 | 7 | 31 | 121 | 91 | 27 | 12 | 2 | 2 | 4 | 1 | 1 | 0 |
| 2005 | 1 | 25 | 71 | 77 | 48 | 22 | 13 | 4 | 0 | 1 | 0 | 1 |
| 2006 | 0 | 17 | 41 | 53 | 38 | 12 | 7 | 1 | 1 | 0 | 2 | 0 |
| 2007 | 0 | 47 | 136 | 61 | 22 | 17 | 4 | 2 | 0 | 0 | 0 | 0 |
| 2008 | 1 | 55 | 106 | 70 | 21 | 5 | 2 | 1 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 13 | 105 | 73 | 28 | 10 | 4 | 0 | 1 | 0 | 0 | 0 |

Table 7.11.4. Catch weight-at-age for plaice in VIIjk.

| $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1993 |  | 0.197 | 0.256 | 0.306 | 0.417 | 0.582 | 0.750 | 0.933 | 1.159 | 1.534 | 1.969 |  |
| 1994 | 0.046 | 0.222 | 0.302 | 0.368 | 0.460 | 0.563 | 0.708 | 0.871 | 1.031 | 1.307 | 1.373 |  |
| 1995 | 0.100 | 0.228 | 0.272 | 0.325 | 0.390 | 0.519 | 0.645 | 0.818 | 1.197 | 1.475 | 1.558 |  |
| 1996 | 0.029 | 0.298 | 0.379 | 0.431 | 0.463 | 0.512 | 0.528 | 0.494 | 0.595 | 2.322 |  |  |
| 1997 | 1.111 | 0.285 | 0.338 | 0.431 | 0.485 | 0.653 | 0.807 | 0.928 |  |  |  | 1.314 |
| 1998 |  | 0.249 | 0.308 | 0.419 | 0.529 | 0.690 | 0.779 | 0.757 | 0.941 | 1.192 | 2.201 |  |
| 1999 | 0.218 | 0.289 | 0.354 | 0.417 | 0.596 | 0.627 | 0.840 | 0.881 | 1.170 | 1.731 | 2.121 | 1.135 |
| 2000 | 0.119 | 0.274 | 0.348 | 0.420 | 0.486 | 0.610 | 0.805 | 1.113 | 1.437 | 1.088 |  | 1.737 |
| 2001 | 0.214 | 0.243 | 0.325 | 0.405 | 0.536 | 0.648 | 0.798 | 0.561 | 1.119 |  |  |  |
| 2002 |  | 0.211 | 0.296 | 0.328 | 0.415 | 0.498 | 0.567 | 0.701 | 1.014 | 1.098 |  | 1.532 |
| 2003 |  | 0.274 | 0.356 | 0.402 | 0.482 | 0.575 | 0.737 | 0.881 | 1.048 | 1.872 | 1.257 |  |
| 2004 | 0.128 | 0.258 | 0.309 | 0.341 | 0.448 | 0.550 | 0.633 | 0.635 | 0.900 | 1.137 | 1.328 | 1.803 |
| 2005 | 0.174 | 0.238 | 0.276 | 0.324 | 0.381 | 0.459 | 0.731 | 0.949 |  | 1.222 | 1.534 | 2.020 |
| 2006 |  | 0.272 | 0.319 | 0.370 | 0.438 | 0.520 | 0.794 | 0.895 | 0.792 |  | 1.880 |  |
| 2007 |  | 0.239 | 0.281 | 0.354 | 0.433 | 0.482 | 0.573 | 0.727 | 1.394 | 0.837 | 1.266 |  |
| 2008 | 0.293 | 0.239 | 0.282 | 0.336 | 0.358 | 0.530 | 0.756 | 0.399 | 1.106 | 1.576 |  |  |
| 2009 |  | 0.224 | 0.255 | 0.335 | 0.403 | 0.462 | 0.520 |  | 1.080 |  | 1.393 | 1.138 |

ple-7h-k
Relative landings and number of samples


Figure 7.11.1. Irish Operational landings and sampling levels (number of samples) for plaice in VIIjk by quarter (top), geartype (middle) and ICES Division (bottom). The sampling appears to be representative of the landings.


Figure 7.11.2. Length frequency distribution of the Irish landings of plaice in VIIjk between 1993 and 2000. All gears and quarters combined. Sampling was poor during 2006 and 2007.


Figure 7.11.3. Age distribution of plaice in VIIjk between 1993 and 2009. All gears and quarters combined. The age data for 1995 were considered insufficient and for this year the combined age data for 1993-1996 were used.

Plaice VIIjk
Standardised catch proportions-at-ag


Figure 7.11.4. Standardised catch proportions-at-age for plaice in VIIjk. Grey bubbles represent higher than average catch-at-age and black bubbles represent lower than average catch-at-age.

## Plaice VIljk

Log catch numbers


Figure 7.11.5. Log catch numbers-at-age (ages 4-8).


Figure 7.11.6. Catch curve of plaice in VIIbc. Plaice from the age of 4 appear to be fully selected; the data get quite noisy from the age of 7 onwards.

Plaice VIIjk
Age range 4-7


Figure 7.11.7. Z estimated over pseudo-cohorts as the slope of the $\log$ catch numbers.


Figure 7.11.8 Log catch numbers (standardised by year). Fish appear to be fully selected from the age of 4 .


Figure 7.11.9. Selectivity was modelled by fitting a line through the mean $\log$ standardised catch numbers of ages 4 to 9 to predict the expected catch numbers for ages $\mathbf{1}$ to 3 if these were fully selected. The proportions of observed divided by expected catch number were taken as the 'observed' selectivity. This was then modelled using a logit transformation.


Figure 7.11.10. YPR analysis using the Thompson-Bell approach. Recent estimates of $Z$ were between 0.5 to 1.2 which translates to an F of 0.48 to 1.08 .

### 7.12 Sole in West of Ireland Division VIIb, c

Type of assessment in 2010
No assessment was performed.

### 7.12.1 General

## Stock Identity

Sole in VIIb are mainly caught by Irish vessels on sandy grounds in coastal areas. Sole catches in VIIc are negligible. In VIIb there are two distinct areas where sole are caught: an area to the west of the Aran Islands and an area in the north of VIIb which extends into VIa (the Stags Ground). The landings and lpue of Sole in VIIbc appear to have been more or less stable since the start of the logbooks time-series in 1995 (WD1, WGCSE 2009). It is not known how much exchange there is between sole on the Aran grounds and those on the Stags ground.

### 7.12.2 Data

The nominal landings are given in Table 7.12.1.

Table 7.12.1. Landings of Sole in VIIbc as officially reported to ICES.

| Country | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | - | 25 | 7 | 6 | 3 | 3 | 6 | 9 | 6 | 5 | 9 | 3 | 6 |
| Ireland | 12 | 12 | 19 | 44 | 14 | 16 | 13 | 24 | 47 | 55 | 40 | 17 | 44 |
| Spain | 19 | 16 | 30 | 25 | 1 | - | 11 | 1 | - | - | - | - | - |
| UK - Eng+1 | . | . | . | . | . | . | . | . | . |  | . | . |  |
| UK - Engla | - | - | - | - | - | - | - | - | - | 1 | - | - | - |
| Total | 31 | 53 | 56 | 75 | 18 | 19 | 30 | 34 | 53 | 61 | 49 | 20 | 50 |
| Country | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| France | 8 | 2 | 2 | - | - | 5 | 2 | 1 | 1 | 2 | 2 | 3 | - |
| Ireland | 29 | 39 | 34 | 38 | 41 | 46 | 43 | 59 | 60 | 59 | 52 | 51 | 49 |
| Spain | - | - | - | - | - | - | - | - | - | - | - | - | - |
| UK - Eng+1 | . | . | . | - | - | - | - | - | - | - | - | 1 | - |
| UK - Engla | - | - | 1 | . | . | . | . | . | . | . | . | . | . |
| Total | 37 | 41 | 37 | 38 | 41 | 51 | 45 | 60 | 61 | 61 | 54 | 55 | 49 |
| Unallocated |  |  |  |  |  |  |  | 0 | 9 | -2 | 3 | 0 | 17 |
| Total as esti | nated by | Workin | roup |  |  |  |  | 60 | 70 | 59 | 57 | 55 | 66 |
| Country | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| France |  | 12 | 7 | 14 | 19 | 18 | 7 | 12 | 7 | 6 |  |  |  |
| Ireland | 68 | 65 | 53 | 50 | 50 | 49 | 38 | 31 | 34 | 31 | 31 |  |  |
| Spain | - | - | - | - | - | - | . | . |  |  |  |  |  |
| UK - Eng+1 | - | - | - | - | 0 | - | . | . |  |  |  |  |  |
| UK - Engla | . | . | . | . | . | . | . | . |  |  |  |  |  |
| Total | 68 | 77 | 60 | 64 | 69 | 67 | 45 | 43 | 41 | 37 | 31 |  |  |
| Unallocate | 4 | -9 | 0 | -3 | -5 | 2 | -1 | 0 | , | 3 |  |  |  |
| Total as es | 72 | 68 | 60 | 61 | 64 | 69 | 44 | 43 | 42 | 40 |  |  |  |

### 7.13 Sole in Divisions VIIfg

## Type of assessment in 2010

Update.

## ICES advice applicable to 2009

In the advice for 2009 ICES considered the stock as having full reproductive capacity and being harvested sustainably.

Single-stock exploitation boundaries
Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects

The current fishing mortality (2007) is estimated to be 0.33, which is above the rate expected to lead to high long-term yields and low risk of stock depletion.

Exploitation boundaries in relation to precautionary limits
F should be kept below Fpa, corresponding to landings of less than 1090 tonnes in 2009. This is expected to keep the stock above Bpa.

Conclusion on exploitation boundaries
In the present situation with a stock that is above Bpa and a fishing mortality below Fpa, there is no long-term gain in yield to increase fishing mortality. ICES therefore recommends to limit landings in 2009 to no more than $940 t$.

ICES advice applicable to 2010
In the advice for 2010 ICES considered the stock as having full reproductive capacity and being harvested sustainably.

Single-stock exploitation boundaries
Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects

The current fishing mortality (2008) is estimated to be 0.27, which is slightly above the rate expected to lead to high long-term yields and low risk of stock depletion.

Exploitation boundaries in relation to precautionary limits
F should be kept below Fpa, corresponding to landings of less than 1185 tonnes in 2010. This is expected to keep the stock above Bpa.

Conclusion on exploitation boundaries
ICES advises that there is no long-term gain in yield to increase fishing mortality. ICES therefore recommends limiting landings in 2010 to no more than $920 t$.

### 7.13.1 General

## Stock description and management units



Red Boxes -TA C/Management Areas
Blue Shading - Assessment Area

A TAC is in place for ICES Divisions VIIfg. These Divisions correspond to the stock area. The basis for the stock assessment Area VIIfg is described in detail in the Stock Annex.

## Management applicable to 2009 and 2010

Management of sole in VIIfg is by TAC and technical measures. The agreed TACs in 2009 and 2010 are presented in the text tables below. Technical measures in force for this stock are minimum mesh sizes and minimum landing size ( 24 cm ). National regulations also restricted areas for certain types of vessels.

| Species: Common sole <br> Solea solea | Zone:VIf and VIIg <br> (SOL//7FG.) |  |
| :--- | :--- | :--- |
| Belgium | 621 |  |
| France | 62 |  |
| Ireland | 31 | Analytical TAC <br> United Kingdom <br> EC |
| TAC | 979 |  |
| applicies. 3 of Regulation (EC) No 847/96 |  |  |
| Article 4 of Regulation (EC) No 847/96 |  |  |
| applies. |  |  |
| Article 5(2) of Regulation (EC) No 847/96 |  |  |
| applies. |  |  |

2010 TAC

| Species: | Common sole <br> Solea solea | Zone: | VIIf and VIIg <br> (SOL/7FG.) |
| :--- | :--- | :--- | :--- |
| Belgium | 621 |  |  |
| France | 62 |  |  |
| Ireland | 31 |  |  |
| United Kingdom | 279 |  |  |
| EU | 993 | Analytical TAC |  |
| TAC | 993 |  |  |

Three rectangles in the Celtic Sea (30E4, 31E4 and 32E3) were closed during the first quarter of 2005, and in February-March 2006, 2007, 2008 and 2009. A derogation permitted beam trawlers to fish in March 2005. This derogation was not continued in the years thereafter. The effects of this closure were discussed in WGSSDS and ACFM 2007. No new information was available at the time of the update Working Group.

## Fishery in 2009

The Working Group estimated the total international landings at 790 t in 2009 (Table 7.13.1), which is about $20 \%$ below the 2009 TAC ( 993 t ) and also $9 \%$ below last year's forecast of 869 t .

Early in the time-series officially reported landings included Divisions VIIg-k for some countries and their total was higher than the WG estimate. Since 1999 official landings correspond to Divisions VIIfg, and the total is lower than the Working Group estimate. During the period 2002-2004 the difference between the two estimates was substantial. This was mainly due to area misreporting, which was taken into account in the Working Group estimates. Although no official landings were available from France, estimates of French landings were provided to the Working Group.

### 7.13.2 Data

## Landings

Irish landings submitted to the Working Group for 2008 were revised upward by $1 \%$ to 28 t . The 2008 values for the numbers-at-age were therefore also updated. Total landings now amount to 800 t (Table 7.13.1).

Annual length compositions for 2008 are given by fleet in Table 7.13.2. Length distributions of the total Belgian and UK (England and Wales) landings for the last eleven years are plotted in Figure 7.13.1. Belgium lands a greater proportion of small fish compared to the UK (England and Wales).

Quarterly numbers and weight-at-age data are available for the Belgian and UK landings (approx. $90 \%$ of the total landings). Catch weights-at-age were calculated, weighted by national catch numbers-at-age, and then quadratically smoothed in year (using age $=1.5,2.5$, etc.) and SOP-corrected. For 2009, the quadratic fit used was:

$$
\mathrm{W}(\mathrm{t})=+0.0617+\left(0.0469^{*}(\mathrm{AGE})\right)-\left(0.00009^{*}(\mathrm{AGE})^{2}\right) \quad \mathrm{R} 2=0.98
$$

Further details on raising procedures are given in the Stock Annex.
Stock weights-at-age were the first quarter catch weights of the Belgium and the UK beam trawl fleets and smoothed by fitting a quadratic fit:

$$
\mathrm{W}(\mathrm{t})=-0.0216+\left(0.0704^{*}(\mathrm{AGE})\right)-\left(0.0012^{*}(\mathrm{AGE})^{2}\right) \quad \mathrm{R} 2=0.99
$$

Catch numbers-at-age are given in Table 7.13.3, and weights-at-age in the catch and the stock are given in Tables 7.13.4-5. Age compositions over the last eleven years are plotted in Figure 7.13.2. The standardised catch proportion-at-age is presented in Figure 7.13.3.

UK has provided data this year under the ICES InterCatch format. Belgium, France and Ireland are working to provide data using this format for the next Working Group.

Sampling levels for those countries providing age compositions are given in Table 1.3.1.

## Discards

The available discard data indicate that discarding of sole is usually minor. In 2007, 2008 and 2009, discarding of sole in the UK fleet was estimated at about $3 \%, 1 \%$ and $6 \%$ respectively in numbers. Discard rates of sole in the Belgian beam trawl fleet were available to the Working Group in 2004-2005, and were about $5 \%$ of the total sole catches. Length distributions of retained and discarded catches of sole for 2007, 2008 and 2009 from samples taken onboard UK vessels are given in Figure 7.13.4.

## Biological

Natural mortality was assumed to be 0.1 for all ages and years. The maturity ogive is based on samples taken during the UK (E\&W) beam-trawl survey of March 1993 and 1994 and is applied to all years of the assessment (See also Stock Annex).

The proportion of M and F before spawning was set to zero.

## Surveys

Standardised abundance indices for the UK beam trawl survey (UK (BTS-3Q)) are shown in Table 7.13.6 and Figure 7.13.5. Abundance-at-age 0 is highly variable and not used in the assessment. The UK survey appears to track the stronger year classes reasonably well for most ages. The internal consistency plot also indicates a reasonable fit for most of the age range (Figure 7.13.6).

## Commercial Ipue

Available estimates of effort and lpue are presented in Tables 7.13.7-8 and Figure 7.13.7.

Belgian beam trawl (BEL-BEAM) effort was at highest levels in 2004-2005. During these years effort shifted from the Eastern English Channel (VIId) to the Celtic Sea because of days at sea limitations in the former area. In 2006, these restrictions had been lifted and effort decreased back to similar levels compared to the early 2000s. The sharp effort reduction in 2008 may be a combined result of the unrestricted effort regime in VIId and the high fuel prices. Effort stayed at the same level in 2009. Lpue peaked in 2002. After a sharp decline to its record low in 2004, lpue has been increasing gradually to average levels of the time-series.

The effort from the UK (E\&W) beam trawl fleet (UK-CBT) has declined sharply since the early 2000s to a record low in 2009. Lpue in the 1990s and 2000s was stable, but at lower levels compared to the period before. In 2007, lpue increased considerably and gave a similar value for 2008. In 2009 there was a decrease to a level just above the mean of the time-series.

Irish effort and lpue data are also presented. The main target species in the Irish fisheries are megrim, anglerfish, etc. The vessels usually operate on fishing grounds in the Western Celtic Sea with lower sole densities.

The internal consistency plots for the main two commercial lpue series, used in the assessment (UK-CBT and BEL-CBT), show high consistencies for the entire age range (Figures 7.13.8-9).

## Other relevant data

Reports from UK industry suggest that the main issues affecting the fishery in VIIfg were displacement of effort due to the rectangle closures and the restrictions on the use of 80 mm mesh west of $7^{\circ} \mathrm{W}$ (Trebilcock and Rozarieux, 2009)

No additional information was received from the Belgian, French or Irish industries.

### 7.13.3 Stock assessment

The method used to assess Celtic Sea sole is XSA, using one survey and two commercial tuning-series (Table 7.13.9). It should be noted that the year range of the Belgian commercial beam trawl tuning fleet only covers 1971 up to 2003 (see also Section 7.13.9 recommendation for next Benchmark). Table 7.13 .9 also includes tuning indices of the Irish ground fish survey (IR-GFS) and the commercial UK otter trawl fleet (UKCOT) which are not used in this assessment.

## Data screening

Adding the 2n009 data to the different time-series, together with the Irish landings revisions for 2008 did not cause any additional anomalies compared to previous
years. The "single fleet runs", "separable VPA", etc. that are used to screen the data of this stock are therefore not presented in this report, but are available in the 'Exploratory runs folder'. This folder also contains a comparison plot of SSB, R and F of last year's final assessment and of the same assessment but with the Irish landings revisions. The output was very similar for both assessments.

The catchability residuals for the final XSA are shown in Figure 7.13 .10 and the XSA tuning diagnostics are given in Table 7.13.10. There is a marked change in the catchability residuals year effect in 2007, 2008 and 2009 for the UK beam trawl fleet (UK-CBT, positive residuals) and for the UK beam trawl survey (UK (BTS-3Q), negative residuals), indicating a conflicting signal between these two fleets.

In this year's assessment the estimates for the recruiting year class 2008 were estimated solely by the UK beam trawl survey UK (BTS-3Q) (Figure 7.13.11). The survivor estimates of the two prominent fleets (the UK (BTS-3Q) survey and the UK-CBT commercial fleet) which have at least $90 \%$ of the weighting for all the ages, differ from each other for most of the ages. The Working Group was not able to clarify that particular issue. The different estimates from the two fleets do not generate a retrospective bias and therefore probably balance off each other in the assessment. The Working Group also assumed that the Trevose closure, a change in special distribution of the UK beam trawl fleet and the ending of the Belgian tuning-series in 2003, may have an influence on the divergence in survivor estimates from both dominant tuning-series.

F shrinkage has a low weighting for all ages ( $<4 \%$ ). The weighting of the survey decreases for the older ages as the commercial UK-CBT fleet is given more weight (Figure 7.13.11).

## Final update assessment

The final settings used in this year's assessment (and since 2006) are as detailed below:

|  | 2010 assessment |  |  |
| :--- | :---: | :---: | :---: |
| Fleets | Years | Ages | $\alpha-\beta$ |
| BEL-CBT commercial | $71-03$ | $2-9$ | $0-1$ |
| UK-CBT commercial | $91-09$ | $2-9$ | $0-1$ |
| UK(BTS-3Q) survey | $88-09$ | $1-9$ | $0.75-0.85$ |
|  |  |  |  |
| -First data year | 1971 |  |  |
| -Last data year | 2009 |  |  |
| -First age | 1 |  |  |
| -Last age | $10+$ |  |  |
| Time-series weights | None |  |  |
| -Model | Mean q model all ages |  |  |
| -Q plateau set at age | 7 |  |  |
| -Survivors estimates shrunk towards mean F | 5 years/5 ages |  |  |
| -s.e. of the means | 1.5 |  |  |
| -Min s.e. for pop. Estimates | 0.3 |  |  |
| -Prior weighting | None |  |  |

Retrospective patterns for the final run are shown in Figure 7.13.12. SSB is generally underestimated and fishing mortality overestimated.

The final XSA output is given in Table 7.13 .11 (fishing mortalities) and Table 7.13.12 (stock numbers). A summary of the XSA results is given in Table 7.13.13 and trends in yield, fishing mortality, recruitment and spawning-stock biomass are shown in Figure 7.13.13.

## Comparison with previous assessment

Figure 7.13.14 gives the historical performance of this stock. The trends in SSB, F and recruitment are consistent from year to year. However, some major revisions in the estimates were made in the period just after 1998. The underlying causes were the exceptionally strong 1998 year class, and the use of a power model in the assessment at that time, which substantially revised the year class.

With the addition of the 2009 data, estimates of fishing mortality and SSB for the most recent years were revised slightly. For example, last year fishing mortality and SSB in 2008 were estimated to be 0.27 and 3128 t . In this year's assessment, the 2008 estimates have been revised downwards by $10 \%$ (fishing mortality) and upwards by $10 \%$ (SSB). The estimated recruitment by XSA in 2008 was revised downward by $32 \%$, however the value used in the forecast (7700 thousand fish) was revised upward by $30 \%$ in this year's assessment.

## State of the stock

Trends in landings, SSB, F (4-8) and recruitment are presented Table 7.13.13 and Figure 7.13.13.

During the eighties fishing mortality increased for this stock. In the following decades fishing mortality fluctuated around this higher level. However fishing mortality has decreased since the late 1990s and was estimated to be 0.19 in 2009, which is very close to the record low value of the time-series.

Recruitment has fluctuated around 5 million recruits with occasional strong year classes. The 1998 year class is estimated to be the strongest in the time-series. The 2007 year class is confirmed by this year's assessment to be the second highest for this stock and the incoming recruitment (year class 2008) is estimated to be above average.

SSB has declined almost continuously from the highest value of 8000 t in 1971 to the lowest observed in the time-series in 1998. The exceptional year class of 1998 has increased SSB to above the long-term average. The two good recruitments in 2007 and 2008 are predicted to keep SSB well above $B_{p a}$.

### 7.13.4 Short-term projections

The 2007 year class was estimated to be around 10.0 million fish at age 1 , which is the second highest value in the time-series and about $30 \%$ lower than estimated last year. The XSA survivor estimate for this year class was used for further prediction.

The 2008 year class in 2009 was estimated by XSA to be 7.3 million one year olds which is above average. The estimates solely coming from the UK (BTS-3Q) survey. The XSA survivor estimates for this year class were used for further prediction.

The long-term GM71-07 recruitment ( 5.0 million) was assumed for the 2009 and subsequent year classes.

The Working Group estimates of year-class strength used for prediction can be summarised as follows:

| Year class | At age in 2010 | XSA | Source |  |
| :---: | :---: | :---: | :--- | :--- |
| 2007 | 3 | 7615 | XSA |  |
| 2008 | 2 | 6624 |  | XSA |
| 2009 | 1 | - | 4998 | GM 1971-07 |
| $2010 \& 2011$ | recruits | - | 4998 | GM 1971-07 |

Population numbers at the start of 2010, estimated for ages 4 and older, were taken from the XSA output.

Fishing mortality was set as the mean over the last three years rescaled to 2009. Weights-at-age in the catch and in the stock are averages for the years 2007-2009. Input data are shown in Table 7.13.14. Results are presented in Table 7.13 .15 (management options) and Table 7.13.16 (detailed output).

Assuming status quo F, implies a catch in 2010 of around 870 t (the agreed TAC is 993 t ) and a catch of 947 t in 2011. Assuming status quo F will result in a SSB of 5050 t in 2011 and 5390 t in 2012.

Assuming status quo F , the proportional contributions of recent year classes to the predicted landings and SSB are given in Table 7.13.17. The assumed GM recruitment accounts for about $4 \%$ of the landings in 2011 and about $8 \%$ of the 2012 SSB.

There are no known specific environmental drivers known for this stock.

### 7.13.5 MSY explorations

Yield-per-recruit results, long-term yield and SSB, conditional on the present exploitation pattern and assuming status quo F in 2009, are given in Table 7.13.18 and Figure 7.13.15. $\mathrm{F}_{\max }$ is estimated to be 0.30 . Long-term yield and SSB (using GM recruitment and $\mathrm{F}_{\text {sq }}$ ) are estimated to be 910 t and 5100 t respectively.

Investigations for possible $\mathrm{F}_{\text {msy }}$ candidates for this stock were done with the PLOTMSY program. The inputs are the standard SEN and SUM files, used to produce the standard graphs (Table 7.13.14). The results are shown in Table 7.13 .19 and Figures 7.13.16-19. The Working Group decided that the use of a "Ricker" and a "smooth hockey stick" were possible candidates as a stock-recruitment relationship for this stock in estimating Fmsy (Figures 7.13.17 and 7.13.18). The "Ricker" was finally chosen as it modelled the stock and recruit estimates somewhat better than the "smooth hockey stick". The analysis also show that $\mathrm{F}_{\max }$ is poorly defined (Figure 7.13.19) and that $\mathrm{F}_{\text {msy }}$ candidates at or below 0.31 may be appropriate for sole in VIIfg.

### 7.13.6 Biological reference points

The Working Group's current approach to reference points is outlined in Section 1.4.4. Current biological reference points are given in the text table below:

| Reference points | ACFM 98 onwards |
| :--- | :--- |
| Flim | 0.52 (based on Floss, WG98) |
| Fpa | 0.37 (Flim x 0.72) |
| Blim | Not defined |
| Bpa | 2200 t (based on Bloss (1991), WG98) |

### 7.13.7 Management plans

There are no explicit management plans for Celtic Sea sole.
In 2006, the Working Group presented results from a series of medium-term scenarios, carried out in conjunction with VIIfg plaice, to simulate some possible management plans for the two stocks Results indicated that an F in the range 0.27 to 0.49 in the long-term would maintain yield at or above $95 \%$ of that given by $\mathrm{F}_{\text {max, }}$ whilst posing a low probability ( $<5 \%$ ) of SSB falling below $\mathrm{B}_{\mathrm{pa}}$. Three year average exploitation patterns were calculated and are given in Figure 7.13.20. The results suggest that the results of the analysis carried out in 2006 can still be used. The results of the $\mathrm{F}_{\text {msy }}$ analysis, carried out during this year's Working Group also confirm that a fishing mortality of 0.31 could be a candidate for a long-term management objective for sole in VIIfg.

### 7.13.8 Uncertainties and bias in assessment and forecast

## Sampling

The major fleets fishing for VIIfg sole are sampled. Sampling is considered to be at a reasonable level (Table 1.3.1).

## Discards

Discard estimates, which are low (Figure 7.13.4) are not included in the assessment.

## Surveys

The UK (BTS-3Q) survey, which is solely responsible for the recruiting estimates, has been able to track year-class strength rather well in the past. However, strong year classes have been revised downward in previous assessments and therefore estimates of the very strong year classes may cause possible bias, especially in the forecast.

## Consistency

Figure 7.13 .14 gives the historical performance of this stock. The trends in SSB, F and $R$ are consistent from year to year. However, some major revisions in the estimates were made in the period just after 1998. The underlying causes were the exceptionally strong 1998 year class, and the use of a power model in the assessment at that time, which substantially revised the year class

With the addition of the 2009 data, estimates of fishing mortality and SSB for the most recent years were revised slightly. Last year fishing mortality and SSB in 2008 were estimated to be 0.27 and 3128 t . In this year's assessment, the 2008 estimates have been revised downwards by $10 \%$ (fishing mortality) and upwards by $10 \%$ (SSB). The estimated recruitment by XSA in 2008 was revised downward by $32 \%$, however the value used in the forecast (7700 thousand fish) was revised upward by $30 \%$ in this year's assessment.

## Misreporting

Area misreporting is known to have been considerable over the period 2002-2004. This was due to a combination of the good 1998 year class still being an important part of the catch composition and restrictive TACs. The area misreporting has been corrected for the years 2002-2006 (method explained in the Report of WGSSDS 2007). Since 2007 the area misreporting that could be estimated was negligible (see Stock Annex).

### 7.13.9 Recommendation for next Benchmark

| Year | Candidate Stock | Supporting Justification | Suggested time | Indicate expertise necessary at benchmark meeting |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | VIIf,g sole | A need to update the Belgian commercial tuning-series. The Belgian beam trawl tuningseries is only used up to 2003, mainly because the estimation of the corresponding lpue series could not be calculated correctly. At the 2009 WKFLAT a possible way of calculating Belgian beam trawl lpue for Division VIId was proposed, using a more realistic horsepower correction method. The proposed method should be investigated, not only for the Belgian beam trawl lpue but also for the UK beam trawl lpue in Division VIIfg, which are the two commercial fleets used in this assessment. <br> A need to investigate the spatial distribution of the major Celtic sea fleets and possible impacts of the Trevose closure. | 2012 | Expert Group members |

### 7.13.10 Management considerations

There is no apparent stock-recruitment relationship for this stock and no evidence of reduced recruitment at low levels of SSB.

SSB has declined almost continuously from the highest value of 8000 t in 1971 to the lowest observed in the time-series in 1998, increased subsequently due to the strong 1998 year class, to above the long-term average. The two good recruitments in 2007 and 2008 are predicted to keep SSB well above $B_{p a}$.

The Celtic Sea is an area without days at sea limitations for demersal fisheries. In this context and given that many demersal vessels are very mobile, changes in effort measures in areas other than the Celtic Sea, can influence the effort regime in the Celtic Sea (cfr. increased effort in Celtic Sea for Belgian beamers during 2004-2005 when days at sea limitations were in place for the Eastern English Channel).

## References

Trebilcock P. and N. de Rozarieux. 2009. National Federation Fishermen's Organisation Annual Fisheries Reports. Cornish Fish Producers Organisation / Seafood Cornwall Training Ltd, March 2009.

ICES. 2009. Report of the Benchmark and Data Compilation Workshop for Flatfish (WKFLAT 2009), 6-13 February 2009, Copenhagen, Denmark. ICES CM 2009/ACOM:31. 192 pp.

Table 7.13.1 - Celtic Sea Sole (ICES Divisions VIIfg). Official Nominal landings and data used by the Working Group (t)

| Year | Belgium | Denmark | France | Ireland | UK(E.\&W,NI.) | UK(Scotland) | Netherlands | Total-Official | Unallocated | Used by WG |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 1039 * | 2 | 146 | 188 | 611 |  | 3 | 1989 | -389 | 1600 |
| 1987 | 701 * | - | 117 | 9 | 437 | - |  | 1264 | -42 | 1222 |
| 1988 | 705 * | - | 110 | 72 | 317 |  |  | 1204 | -58 | 1146 |
| 1989 | 684 * | - | 87 | 18 | 203 |  |  | 992 | 0 | 992 |
| 1990 | 716 * | - | 130 | 40 | 353 | 0 |  | 1239 | -50 | 1189 |
| 1991 | 982 * | - | 80 | 32 | 402 | 0 | - | 1496 | -389 | 1107 |
| 1992 | 543 * | - | 141 | 45 | 325 | 6 | - | 1060 | -79 | 981 |
| 1993 | 575 * | - | 108 | 51 | 285 | 11 | - | 1030 | -102 | 928 |
| 1994 | 619 * | - | 90 | 37 | 264 | 8 | - | 1018 | -9 | 1009 |
| 1995 | 763 * | - | 88 | 20 | 294 | - | - | 1165 | -8 | 1157 |
| 1996 | 695 * | - | 102 | 19 | 265 | 0 | - | 1081 | -86 | 995 |
| 1997 | 660 * | - | 99 | 28 | 251 | 0 | - | 1038 | -111 | 927 |
| 1998 | 675 * | - | 98 | 42 | 198 | - | - | 1013 | -138 | 875 |
| 1999 | 604 | - | 61 | 51 | 231 | 0 | - | 947 | 65 | 1012 |
| 2000 | 694 | - | 74 | 29 | 243 | - | - | 1040 | 51 | 1091 |
| 2001 | 720 | - | 77 | 35 | 288 | - | - | 1120 | 48 | 1168 |
| 2002 | 703 | - | 65 | 32 | 318 | + | - | 1118 | 227 | 1345 |
| 2003 | 715 | - | 124 | 26 | 342 | + | - | 1207 | 185 | 1392 |
| 2004 | 735 | - | 79 | 33 | 283 | - | - | 1130 | 119 | 1249 |
| 2005 | 645 | - | 101 | 34 | 217 | - | - | 997 | 47 | 1044 |
| 2006 | 576 | - | 75 | 38 | 232 | - | - | 921 | 25 | 946 |
| 2007 | 582 | - | 85 | 32 | 244 | - | - | 943 | 2 | 945 |
| 2008 | 466 | - | 68 | 28 | 218 | - | - | 780 | 20 | 800 |
| $2009{ }^{1}$ | 511 | - | n/a | 27.82 | 194 | - | - | 733 | 57 | 790 |

Table 7.13.2 - Sole in VIIfg. Annual length distributions by fleet

| Length (cm) | UK (England \& Wales) Beam trawl | Belgium <br> All gears | Ireland* <br> All gears |
| :---: | :---: | :---: | :---: |
| 17 |  |  |  |
| 18 |  |  |  |
| 19 |  |  |  |
| 20 |  |  |  |
| 21 |  |  | 2 |
| 22 | 25 | 204 | 9 |
| 23 | 2334 | 110022 | 42 |
| 24 | 8225 | 235141 | 66 |
| 25 | 18692 | 237347 | 95 |
| 26 | 30860 | 217947 | 125 |
| 27 | 37298 | 215261 | 172 |
| 28 | 41288 | 186040 | 175 |
| 29 | 32026 | 116905 | 193 |
| 30 | 36929 | 119425 | 201 |
| 31 | 29516 | 92349 | 175 |
| 32 | 26603 | 90993 | 205 |
| 33 | 23999 | 60246 | 218 |
| 34 | 23095 | 58462 | 188 |
| 35 | 27744 | 57544 | 160 |
| 36 | 19256 | 45686 | 170 |
| 37 | 15726 | 38334 | 124 |
| 38 | 14922 | 33386 | 108 |
| 39 | 14960 | 23506 | 90 |
| 40 | 16514 | 20261 | 54 |
| 41 | 9799 | 15977 | 45 |
| 42 | 6557 | 8256 | 28 |
| 43 | 6447 | 6916 | 14 |
| 44 | 5959 | 1904 | 18 |
| 45 | 3576 | 1361 | 11 |
| 46 | 3196 | 1212 | 5 |
| 47 | 1921 | 272 | 5 |
| 48 | 571 | 445 | 1 |
| 49 | 176 | 74 |  |
| 50 | 24 |  |  |
| 51 | 253 |  |  |
| 52 | 0 |  |  |
| 53 | 665 |  |  |
| 54 |  |  |  |
| 55 |  |  |  |
| 56 |  |  |  |
| 57 |  |  |  |
| 58 |  |  |  |
| 59 |  |  |  |
| 60 |  |  |  |
| Total | 459157 | 1995476 | 2699 |

* Distributions from sample only

Table 7.13.3 - Sole in VIIfg. Catch numbers at age (in thousands)

|  | YEAR | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | 2 | 386 | 541 | 364 | 155 | 119 | 312 | 314 | 318 | 328 |  |
|  | 3 | 270 | 902 | 1882 | 438 | 287 | 834 | 438 | 741 | 560 |  |
|  | 4 | 1341 | 314 | 748 | 863 | 336 | 560 | 349 | 339 | 747 |  |
|  | 5 | 625 | 670 | 305 | 411 | 638 | 611 | 271 | 154 | 208 |  |
|  | 6 | 433 | 329 | 352 | 209 | 304 | 559 | 244 | 159 | 154 |  |
|  | 7 | 537 | 213 | 119 | 239 | 110 | 261 | 404 | 99 | 197 |  |
|  | 8 | 763 | 232 | 110 | 97 | 102 | 131 | 120 | 198 | 124 |  |
|  | 9 | 376 | 314 | 116 | 109 | 67 | 197 | 28 | 71 | 153 |  |
|  | +gp | 1220 | 730 | 644 | 541 | 372 | 463 | 365 | 174 | 169 |  |
|  | TOTALNUN | 5951 | 4245 | 4640 | 3062 | 2335 | 3928 | 2533 | 2253 | 2640 |  |
|  | TONSLAND | 1861 | 1278 | 1391 | 1105 | 919 | 1350 | 961 | 780 | 954 |  |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |  |
|  | YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 |  | 657 | 602 | 342 | 647 | 672 | 196 | 494 | 318 | 526 | 479 |
| 3 |  | 972 | 675 | 831 | 1078 | 846 | 1473 | 1296 | 957 | 464 | 1164 |
| 4 |  | 876 | 792 | 309 | 729 | 606 | 766 | 1173 | 797 | 879 | 601 |
| 5 |  | 584 | 399 | 467 | 284 | 542 | 565 | 526 | 577 | 441 | 621 |
| 6 |  | 180 | 377 | 280 | 349 | 184 | 296 | 358 | 273 | 387 | 237 |
| 7 |  | 62 | 150 | 207 | 225 | 277 | 100 | 193 | 205 | 127 | 188 |
| 8 |  | 96 | 120 | 92 | 192 | 106 | 140 | 87 | 100 | 78 | 82 |
| 9 |  | 100 | 94 | 111 | 52 | 47 | 73 | 103 | 61 | 67 | 24 |
|  | +gp | 352 | 380 | 326 | 320 | 274 | 240 | 328 | 179 | 268 | 102 |
| 0 | TOTALNUN | 3879 | 3589 | 2965 | 3876 | 3554 | 3849 | 4558 | 3467 | 3237 | 3498 |
|  | TONSLANE | 1314 | 1212 | 1128 | 1373 | 1266 | 1328 | 1600 | 1222 | 1146 | 992 |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 |  | 277 | 1458 | 433 | 354 | 295 | 129 | 177 | 245 | 197 | 608 |
| 3 |  | 994 | 690 | 1700 | 863 | 790 | 1156 | 1035 | 890 | 932 | 1718 |
| 4 |  | 1176 | 658 | 644 | 1104 | 739 | 1098 | 904 | 599 | 724 | 834 |
| 5 |  | 399 | 496 | 409 | 332 | 864 | 420 | 424 | 400 | 297 | 282 |
| 6 |  | 452 | 151 | 253 | 186 | 283 | 483 | 229 | 252 | 171 | 143 |
| 7 |  | 138 | 156 | 61 | 161 | 149 | 133 | 192 | 127 | 108 | 80 |
| 8 |  | 115 | 55 | 59 | 63 | 65 | 112 | 57 | 126 | 51 | 31 |
| 9 |  | 50 | 46 | 28 | 83 | 42 | 65 | 43 | 45 | 52 | 23 |
|  | +gp | 129 | 162 | 89 | 99 | 146 | 109 | 106 | 106 | 87 | 44 |
| 0 | TOTALNUN | 3730 | 3872 | 3676 | 3245 | 3373 | 3705 | 3167 | 2790 | 2619 | 3763 |
|  | TONSLANL | 1189 | 1107 | 981 | 928 | 1009 | 1157 | 995 | 927 | 875 | 1012 |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 |  | 1721 | 704 | 29 | 119 | 425 | 271 | 685 | 335 | 211 | 599 |
| 3 |  | 1480 | 1918 | 1465 | 697 | 1721 | 855 | 1330 | 865 | 447 | 458 |
| 4 |  | 683 | 860 | 2202 | 1134 | 792 | 837 | 715 | 743 | 552 | 421 |
| 5 |  | 241 | 436 | 660 | 1860 | 794 | 473 | 576 | 474 | 558 | 343 |
| 6 |  | 60 | 242 | 249 | 402 | 721 | 398 | 163 | 325 | 274 | 289 |
| 7 |  | 56 | 65 | 95 | 223 | 114 | 348 | 148 | 157 | 196 | 172 |
| 8 |  | 43 | 39 | 54 | 80 | 60 | 48 | 178 | 145 | 75 | 102 |
| 9 |  | 19 | 26 | 36 | 26 | 34 | 41 | 44 | 184 | 108 | 43 |
|  | +gp | 51 | 81 | 51 | 75 | 49 | 43 | 51 | 70 | 171 | 190 |
| 0 | TOTALNUN | 4354 | 4371 | 4841 | 4616 | 4710 | 3314 | 3890 | 3298 | 2592 | 2617 |
|  | TONSLANL | 1091 | 1168 | 1345 | 1392 | 1249 | 1044 | 946 | 945 | 800 | 790 |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table 7.13.4 - Sole in VIIfg. Catch weights at age (kg)
Run title : CELTIC SEA SOLE,2010WG,COMBSEX,PLUSGROUP At 21/04/2010 17:41

|  | YEAR | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0.039 | 0.106 | 0.081 | 0.063 | 0.046 | 0.114 | 0.098 | 0.068 | 0.023 |  |
| 2 |  | 0.106 | 0.147 | 0.143 | 0.137 | 0.132 | 0.167 | 0.169 | 0.154 | 0.132 |  |
| 3 |  | 0.167 | 0.186 | 0.202 | 0.205 | 0.212 | 0.218 | 0.235 | 0.234 | 0.232 |  |
| 4 |  | 0.222 | 0.226 | 0.258 | 0.270 | 0.286 | 0.268 | 0.297 | 0.309 | 0.321 |  |
| 5 |  | 0.272 | 0.264 | 0.311 | 0.329 | 0.355 | 0.316 | 0.355 | 0.378 | 0.401 |  |
| 6 |  | 0.315 | 0.302 | 0.361 | 0.385 | 0.417 | 0.363 | 0.409 | 0.441 | 0.471 |  |
| 7 |  | 0.352 | 0.340 | 0.408 | 0.436 | 0.473 | 0.409 | 0.460 | 0.499 | 0.531 |  |
| 8 |  | 0.383 | 0.376 | 0.452 | 0.483 | 0.523 | 0.453 | 0.506 | 0.551 | 0.581 |  |
| 9 |  | 0.408 | 0.413 | 0.493 | 0.525 | 0.567 | 0.496 | 0.548 | 0.598 | 0.622 |  |
|  | +gp | 0.4397 | 0.5384 | 0.6021 | 0.6239 | 0.6715 | 0.6649 | 0.6681 | 0.7196 | 0.6636 |  |
|  | YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0.048 | 0.078 | 0.061 | 0.085 | 0.019 | 0.089 | 0.046 | 0.048 | 0.074 | 0.013 |
| 2 |  | 0.144 | 0.154 | 0.156 | 0.173 | 0.131 | 0.170 | 0.144 | 0.146 | 0.157 | 0.109 |
| 3 |  | 0.234 | 0.225 | 0.243 | 0.255 | 0.235 | 0.246 | 0.236 | 0.236 | 0.235 | 0.198 |
| 4 |  | 0.316 | 0.292 | 0.324 | 0.330 | 0.330 | 0.317 | 0.321 | 0.320 | 0.309 | 0.280 |
| 5 |  | 0.392 | 0.355 | 0.397 | 0.398 | 0.416 | 0.383 | 0.400 | 0.396 | 0.378 | 0.355 |
| 6 |  | 0.461 | 0.414 | 0.462 | 0.459 | 0.494 | 0.444 | 0.471 | 0.466 | 0.442 | 0.424 |
| 7 |  | 0.523 | 0.469 | 0.521 | 0.514 | 0.562 | 0.500 | 0.536 | 0.528 | 0.502 | 0.487 |
| 8 |  | 0.579 | 0.519 | 0.572 | 0.561 | 0.622 | 0.552 | 0.594 | 0.584 | 0.557 | 0.543 |
| 9 |  | 0.627 | 0.565 | 0.617 | 0.602 | 0.673 | 0.598 | 0.645 | 0.632 | 0.608 | 0.592 |
|  | +gp | 0.720 | 0.665 | 0.704 | 0.679 | 0.772 | 0.703 | 0.748 | 0.740 | 0.739 | 0.691 |
| 0 | SOPCOFA | 0.999 | 1.000 | 0.999 | 1.000 | 0.999 | 1.002 | 1.000 | 1.001 | 0.999 | 0.999 |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0.049 | 0.054 | 0.073 | 0.057 | 0.081 | 0.068 | 0.027 | 0.074 | 0.079 | 0.015 |
| 2 |  | 0.134 | 0.150 | 0.147 | 0.134 | 0.151 | 0.147 | 0.124 | 0.156 | 0.163 | 0.122 |
| 3 |  | 0.214 | 0.239 | 0.216 | 0.207 | 0.216 | 0.220 | 0.214 | 0.234 | 0.244 | 0.222 |
| 4 |  | 0.291 | 0.320 | 0.281 | 0.275 | 0.276 | 0.288 | 0.296 | 0.307 | 0.320 | 0.315 |
| 5 |  | 0.363 | 0.393 | 0.342 | 0.338 | 0.331 | 0.351 | 0.372 | 0.376 | 0.393 | 0.400 |
| 6 |  | 0.43 | 0.459 | 0.398 | 0.396 | 0.38 | 0.409 | 0.439 | 0.44 | 0.462 | 0.478 |
| 7 |  | 0.494 | 0.516 | 0.451 | 0.45 | 0.425 | 0.462 | 0.5 | 0.5 | 0.528 | 0.549 |
| 8 |  | 0.553 | 0.566 | 0.499 | 0.500 | 0.465 | 0.510 | 0.552 | 0.555 | 0.589 | 0.613 |
| 9 |  | 0.609 | 0.608 | 0.543 | 0.545 | 0.500 | 0.553 | 0.598 | 0.605 | 0.647 | 0.670 |
|  | +gp | 0.747 | 0.674 | 0.640 | 0.645 | 0.563 | 0.643 | 0.677 | 0.707 | 0.781 | 0.766 |
| 0 | SOPCOFA | 0.999 | 1.000 | 1.000 | 0.999 | 1.000 | 0.998 | 1.001 | 1.000 | 0.999 | 1.001 |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 1 |  | 0.078 | 0.066 | 0.054 | 0.123 | 0.066 | 0.068 | 0.085 | 0.075 | 0.098 | 0.132 |
| 2 |  | 0.166 | 0.148 | 0.130 | 0.171 | 0.130 | 0.145 | 0.139 | 0.139 | 0.155 | 0.178 |
| 3 |  | 0.248 | 0.225 | 0.202 | 0.218 | 0.194 | 0.219 | 0.192 | 0.200 | 0.209 | 0.225 |
| 4 |  | 0.322 | 0.296 | 0.271 | 0.266 | 0.256 | 0.288 | 0.245 | 0.258 | 0.26 | 0.271 |
| 5 |  | 0.39 | 0.363 | 0.336 | 0.313 | 0.317 | 0.354 | 0.297 | 0.313 | 0.31 | 0.317 |
| 6 |  | 0.451 | 0.425 | 0.399 | 0.361 | 0.377 | 0.415 | 0.349 | 0.365 | 0.356 | 0.363 |
| 7 |  | 0.506 | 0.482 | 0.457 | 0.408 | 0.435 | 0.473 | 0.4 | 0.414 | 0.401 | 0.408 |
| 8 |  | 0.553 | 0.533 | 0.513 | 0.454 | 0.493 | 0.528 | 0.451 | 0.46 | 0.443 | 0.454 |
| 9 |  | 0.594 | 0.579 | 0.564 | 0.501 | 0.549 | 0.578 | 0.501 | 0.503 | 0.482 | 0.499 |
|  | +gp | 0.6649 | 0.6773 | 0.7045 | 0.6379 | 0.7217 | 0.6918 | 0.6177 | 0.6087 | 0.5448 | 0.6037 |
| 0 | SOPCOFA | 1 | 0.9954 | 1.0001 | 1.0019 | 1.0003 | 1.0004 | 0.9992 | 0.9999 | 1.0035 | 1 |

Table 7.13.5 - Sole in VIIfg. Stock weights at age (kg)
Run title : CELTIC SEA SOLE,2010WG,COMBSEX,PLUSGROUP At 21/04/2010 17:41


Table 7.13.6 - Sole in VIIfg. Indices of abundance (No/100km) for UK(BTS-3Q) survey

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 30 | 81 | 326 | 49 | 19 | 5 | 0 | 0 | 0 | 0 |
| 1989 | 144 | 222 | 331 | 176 | 20 | 15 | 7 | 4 | 2 | 2 |
| 1990 | 30 | 385 | 313 | 50 | 16 | 4 | 7 | 3 | 0 | 0 |
| 1991 | 32 | 241 | 517 | 67 | 17 | 15 | 4 | 0 | 2 | 2 |
| 1992 | 4 | 394 | 260 | 139 | 30 | 18 | 10 | 1 | 2 | 1 |
| 1993 | 3 | 169 | 320 | 43 | 19 | 1 | 2 | 2 | 1 | 1 |
| 1994 | 1 | 333 | 387 | 99 | 14 | 7 | 7 | 0 | 0 | 2 |
| 1995 | 27 | 124 | 222 | 52 | 11 | 6 | 12 | 1 | 1 | 1 |
| 1996 | 3 | 150 | 211 | 54 | 23 | 6 | 2 | 3 | 1 | 2 |
| 1997 | 32 | 433 | 180 | 18 | 11 | 12 | 4 | 3 | 5 | 0 |
| 1998 | 90 | 770 | 411 | 50 | 9 | 7 | 4 | 2 | 1 | 5 |
| 1999 | 24 | 2464 | 250 | 32 | 14 | 5 | 4 | 4 | 1 | 0 |
| 2000 | 13 | 916 | 1356 | 31 | 22 | 5 | 0 | 2 | 1 | 1 |
| 2001 | 22 | 379 | 599 | 259 | 20 | 7 | 5 | 2 | 0 | 2 |
| 2002 | 8 | 663 | 238 | 127 | 102 | 12 | 6 | 2 | 3 | 0 |
| 2003 | 12 | 392 | 530 | 47 | 26 | 47 | 8 | 3 | 3 | 0 |
| 2004 | 55 | 750 | 377 | 87 | 13 | 19 | 37 | 4 | 2 | 0 |
| 2005 | 37 | 343 | 225 | 32 | 14 | 6 | 4 | 14 | 1 | 2 |
| 2006 | 11 | 273 | 201 | 39 | 13 | 7 | 0 | 2 | 10 | 0 |
| 2007 | 88 | 357 | 108 | 43 | 14 | 11 | 6 | 3 | 3 | 12 |
| 2008 | 5 | 1039 | 104 | 13 | 15 | 6 | 8 | 3 | 3 | 4 |
| 2009 | 1 | 509 | 318 | 24 | 6 | 8 | 3 | 2 | 2 | 2 |
| Geomean | 15 | 383 | 300 | 53 | 17 | 8 | 6 | 3 | 2 | 2 |
| Mean | 30 | 518 | 354 | 70 | 20 | 11 | 6 | 3 | 2 | 2 |

Table 7.13.7 - Sole in VIIfg. Indices of effort.

| Year | England \& Wales |  | Belgium |  |  | Ireland |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Otter trawl | Beam trawl ${ }^{1}$ | Beam trawl ${ }^{2}$ | Beam trawl ${ }^{4}$ | Otter trawl ${ }^{3}$ | Scottish seine ${ }^{4}$ | Beam trawl ${ }^{4}$ |
| 1971 |  |  | 11.06 |  |  |  |  |
| 1972 | 45.72 |  | 8.44 |  |  |  |  |
| 1973 | 45.28 |  | 17.39 |  |  |  |  |
| 1974 | 38.94 |  | 18.83 |  |  |  |  |
| 1975 | 33.53 |  | 16.38 |  |  |  |  |
| 1976 | 25.61 |  | 28.07 |  |  |  |  |
| 1977 | 27.16 |  | 24.11 |  |  |  |  |
| 1978 | 27.08 |  | 18.09 |  |  |  |  |
| 1979 | 23.84 |  | 18.90 |  |  |  |  |
| 1980 | 26.43 |  | 29.02 |  |  |  |  |
| 1981 | 24.10 |  | 35.39 |  |  |  |  |
| 1982 | 19.20 |  | 28.77 |  |  |  |  |
| 1983 | 17.61 |  | 34.95 |  |  |  |  |
| 1984 | 23.16 |  | 33.48 |  |  |  |  |
| 1985 | 25.24 | 18.70 | 40.49 |  |  |  |  |
| 1986 | 21.18 | 20.72 | 52.46 |  |  |  |  |
| 1987 | 24.43 | 38.76 | 37.26 |  |  |  |  |
| 1988 | 20.09 | 25.62 | 42.92 |  |  |  |  |
| 1989 | 17.61 | 20.26 | 53.58 |  |  |  |  |
| 1990 | 22.56 | 30.77 | 40.27 |  |  |  |  |
| 1991 | 18.57 | 40.81 | 18.05 |  |  |  |  |
| 1992 | 16.00 | 35.78 | 25.47 |  |  |  |  |
| 1993 | 13.79 | 39.64 | 31.27 |  |  |  |  |
| 1994 | 9.48 | 37.03 | 38.35 |  |  |  |  |
| 1995 | 8.46 | 37.59 | 47.81 |  | 63.56 | 6.43 | 20.78 |
| 1996 | 8.67 | 39.78 | 47.63 | 53.27 | 60.04 | 9.73 | 26.76 |
| 1997 | 8.14 | 43.00 | 51.98 | 57.36 | 65.10 | 16.13 | 28.25 |
| 1998 | 7.13 | 47.84 | 52.11 | 57.79 | 72.30 | 14.94 | 35.25 |
| 1999 | 5.69 | 50.87 | 55.03 | 55.11 | 51.66 | 8.01 | 40.87 |
| 2000 | 4.05 | 51.19 | 56.05 | 51.34 | 60.60 | 9.90 | 37.03 |
| 2001 | 4.42 | 49.32 | 52.06 | 54.90 | 69.43 | 16.33 | 39.71 |
| 2002 | 6.10 | 37.53 | 43.24 | 49.60 | 77.69 | 20.86 | 31.62 |
| 2003 | 9.94 | 40.71 | 42.81 | 62.73 | 86.79 | 20.91 | 49.26 |
| 2004 | 9.42 | 32.37 |  | 78.73 | 96.99 | 19.38 | 54.86 |
| 2005 | 12.09 | 27.73 |  | 64.50 | 124.40 | 14.81 | 49.65 |
| 2006 | 12.97 | 18.57 |  | 50.28 | 119.23 | 14.79 | 60.48 |
| 2007 | 10.66 | 15.37 |  | 45.72 | 136.53 | 15.82 | 55.86 |
| 2008 | 10.13 | 13.83 |  | 28.71 | 125.81 | 11.65 | 37.22 |
| 2009 | 8.97 | 12.23 |  | 30.85 | 135.18 | 8.15 | 37.92 |

${ }^{1}$ Division VIIf only - Fishing hours (x10^3) corrected for fishing power
${ }^{2}$ Fishing hours ( $\mathrm{x} 10^{\wedge} 3$ ) corrected for fishing power using $\mathrm{P}=0.000204 \mathrm{BHP}^{\wedge} 1.23$
${ }^{3}$ Division VIIg only - Fishing hours ( $\times 10^{\wedge} 3$ )
${ }^{4}$ Fishing hours ( $\mathrm{x} 10^{\wedge} 3$ )

Table 7.13.8 - Sole in VIIfg. LPUE

| Year | UK | England \& Wales |  |  | Belgium |  | Ireland |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | BT Survey ${ }^{4}$ | Otter trawl ${ }^{1}$ | Otter trawl ${ }^{1}$ | Beam trawl ${ }^{1}$ | Beam trawl ${ }^{2}$ | Beam trawl ${ }^{5}$ | Otter trawl ${ }^{5}$ | Scottish sein ${ }^{5}$ | Beam trawl ${ }^{5}$ |
|  | Division Vllfg | Division VIIf | Division VIIg ${ }^{3}$ | Division VIIf | Division VIIfg | Division VIIfg | Division VIIg | Division VIIg | Division VIIg |
| 1971 | - |  |  | - | 47.92 |  |  |  |  |
| 1972 | - | 2.42 | 2.11 | - | 37.06 |  |  |  |  |
| 1973 | - | 2.45 | 0.98 | - | 39.47 |  |  |  |  |
| 1974 | - | 2.10 | 1.83 | - | 37.81 |  |  |  |  |
| 1975 | - | 1.82 | 1.79 | - | 31.41 |  |  |  |  |
| 1976 | - | 2.02 | 1.30 | - | 30.50 |  |  |  |  |
| 1977 | - | 1.84 | 1.21 | - | 27.90 |  |  |  |  |
| 1978 | - | 1.82 | 1.17 |  | 23.35 |  |  |  |  |
| 1979 | - | 1.80 | 1.15 |  | 33.19 |  |  |  |  |
| 1980 | - | 1.86 | 1.55 |  | 29.73 |  |  |  |  |
| 1981 | - | 1.45 | 0.60 |  | 24.03 |  |  |  |  |
| 1982 | - | 1.73 | 0.56 |  | 25.93 |  |  |  |  |
| 1983 | - | 2.22 | 1.14 |  | 22.18 |  |  |  |  |
| 1984 | - | 1.53 | 1.70 |  | 20.78 |  |  |  |  |
| 1985 | - | 1.55 | 1.55 | 12.52 | 17.94 |  |  |  |  |
| 1986 | - | 1.38 | 0.99 | 10.94 | 17.83 |  |  |  |  |
| 1987 | - | 0.94 | 1.15 | 7.31 | 17.32 |  |  |  |  |
| 1988 | 71.14 | 0.62 | 0.27 | 4.39 | 15.29 |  |  |  |  |
| 1989 | 135.18 | 0.99 | 0.87 | 5.38 | 11.33 |  |  |  |  |
| 1990 | 90.67 | 0.76 | 0.67 | 5.98 | 15.64 |  |  |  |  |
| 1991 | 122.88 | 0.69 | 0.85 | 4.80 | 24.24 |  |  |  |  |
| 1992 | 115.79 | 1.00 | 1.25 | 4.14 | 18.57 |  |  |  |  |
| 1993 | 75.42 | 0.55 | 0.25 | 4.80 | 15.21 |  |  |  |  |
| 1994 | 107.77 | 0.90 | 0.27 | 4.26 | 13.94 |  |  |  |  |
| 1995 | 72.50 | 0.96 | 0.87 | 4.52 | 13.62 |  | 0.40 | 0.62 | 0.81 |
| 1996 | 70.15 | 0.66 | 0.52 | 3.94 | 11.27 | 11.45 | 0.73 | 0.05 | 0.88 |
| 1997 | 81.66 | 0.86 | 0.52 | 3.28 | 9.96 | 9.68 | 0.42 | 0.23 | 1.16 |
| 1998 | 135.41 | 0.60 | 0.40 | 2.67 | 10.12 | 9.64 | 0.48 | 0.11 | 1.11 |
| 1999 | 168.46 | 0.91 | 0.74 | 3.21 | 11.26 | 12.14 | 0.17 | 0.09 | 0.50 |
| 2000 | 236.43 | 0.49 | 1.85 | 3.36 | 11.90 | 13.77 | 0.19 | 0.05 | 0.26 |
| 2001 | 154.79 | 1.14 | 2.13 | 4.02 | 13.25 | 13.60 | 0.27 | 0.55 | 0.15 |
| 2002 | 118.11 | 0.78 | 3.60 | 5.64 | 18.71 | 17.80 | 0.43 | 0.29 | 0.14 |
| 2003 | 123.93 | 0.57 | 0.00 | 5.23 | 19.48 | 11.40 | 0.12 | 0.03 | 0.20 |
| 2004 | 149.65 | 0.60 | 0.19 | 5.75 |  | 9.17 | 0.18 | 0.02 | 0.20 |
| 2005 | 76.26 | 0.76 | 0.26 | 4.94 |  | 9.78 | 0.14 |  | 0.28 |
| 2006 | 68.96 | 1.16 | 0.60 | 5.97 |  | 10.70 | 0.11 | 0.05 | 0.26 |
| 2007 | 80.95 | 0.78 | 1.00 | 9.87 |  | 11.74 | 0.13 | 0.02 | 0.20 |
| 2008 | 115.96 | 0.82 | 0.86 | 9.46 |  | 14.51 | 0.12 | 0.02 | 0.29 |
| 2009 | 89.80 | 0.94 | 0.46 | 6.61 |  | 12.90 | 0.10 | 0.00 | 0.28 |

${ }^{1} \mathrm{Kg} / \mathrm{hr}$ corrected for GRT.
${ }^{2} \mathrm{Kg} / \mathrm{hr}$ corrected for fishing power using $\mathrm{P}=0.000204 \mathrm{BH} \mathrm{P}^{\wedge} 1.23$
${ }^{3}$ Division VIIg (East).
${ }^{4} \mathrm{Kg} / 100 \mathrm{~km}$
${ }^{5} \mathrm{Kg} /$ hour

Table 7.13.9-Sole in Vllfg. Tuning series
Indices in bold are used in the assessment

| BEL-CBT |  | Belgium Beam trawl (Effort = Corrected formula) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 197112 | 2003 |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
|  |  | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 11.06 |  | 111 | 77 | 384 | 179 | 124 | 154 | 218 | 108 | 32 | 107 | 76 | 21 | 40 |
| 8.44 |  | 132 | 220 | 76 | 163 | 80 | 52 | 57 | 76 | 39 | 23 | 14 | 38 | 14 |
| 17.39 |  | 179 | 926 | 368 | 150 | 173 | 58 | 54 | 57 | 108 | 32 | 23 | 21 | 45 |
| 18.83 |  | 102 | 287 | 565 | 270 | 136 | 156 | 64 | 79 | 90 | 75 | 38 | 39 | 37 |
| 16.38 |  | 69 | 167 | 195 | 370 | 176 | 64 | 59 | 39 | 33 | 29 | 37 | 18 | 23 |
| 28.07 |  | 199 | 533 | 357 | 391 | 357 | 167 | 84 | 125 | 40 | 17 | 21 | 51 | 35 |
| 24.11 |  | 220 | 307 | 244 | 190 | 170 | 283 | 84 | 20 | 35 | 39 | 36 | 18 | 52 |
| 18.09 |  | 173 | 403 | 185 | 84 | 86 | 54 | 108 | 38 | 11 | 21 | 61 | 8 | 9 |
| 18.9 |  | 222 | 379 | 506 | 141 | 104 | 133 | 84 | 103 | 35 | 12 | 16 | 4 | 6 |
| 29.02 |  | 438 | 647 | 583 | 389 | 119 | 45 | 63 | 66 | 92 | 22 | 25 | 16 | 10 |
| 35.39 |  | 429 | 481 | 565 | 286 | 268 | 107 | 86 | 67 | 86 | 74 | 33 | 13 | 13 |
| 28.77 |  | 245 | 594 | 221 | 334 | 200 | 148 | 66 | 80 | 54 | 19 | 41 | 16 | 25 |
| 34.95 |  | 363 | 605 | 409 | 159 | 196 | 127 | 108 | 29 | 44 | 32 | 15 | 12 | 12 |
| 33.48 |  | 372 | 467 | 334 | 300 | 102 | 153 | 59 | 26 | 26 | 16 | 24 | 19 | 18 |
| 40.49 |  | 52 | 909 | 471 | 372 | 208 | 75 | 104 | 46 | 68 | 15 | 29 | 16 | 10 |
| 52.46 |  | 377 | 900 | 823 | 359 | 230 | 140 | 49 | 58 | 65 | 29 | 50 | 6 | 9 |
| 37.23 |  | 247 | 664 | 438 | 344 | 191 | 119 | 47 | 29 | 20 | 4 | 14 | 2 | 16 |
| 42.92 |  | 362 | 293 | 603 | 250 | 197 | 77 | 51 | 36 | 26 | 19 | 19 | 13 | 16 |
| 53.58 |  | 244 | 680 | 428 | 471 | 179 | 145 | 62 | 13 | 24 | 10 | 19 | 3 | 17 |
| 40.27 |  | 231 | 742 | 663 | 181 | 240 | 70 | 59 | 17 | 26 | 12 | 2 | 4 | 12 |
| 18.05 |  | 1028 | 380 | 225 | 131 | 29 | 26 | 9 | 7 | 13 | 8 | 4 | 1 | 2 |
| 25.47 |  | 327 | 1062 | 376 | 210 | 98 | 14 | 14 | 7 | 9 | 5 | 0 | 0.3 | 2 |
| 31.27 |  | 296 | 615 | 629 | 161 | 81 | 75 | 38 | 36 | 19 | 4 | 2 | 1 | 1 |
| 38.35 |  | 205 | 524 | 523 | 530 | 176 | 71 | 20 | 15 | 16 | 11 | 6 | 5 | 7 |
| 47.81 |  | 77 | 827 | 838 | 277 | 250 | 78 | 48 | 21 | 17 | 8 | 1 | 5 | 2 |
| 47.63 |  | 104 | 737 | 579 | 258 | 130 | 88 | 29 | 17 | 9 | 12 | 3 | 3 | 0 |
| 51.98 |  | 193 | 661 | 377 | 241 | 143 | 74 | 55 | 23 | 16 | 18 | 7 | 3 | 2 |
| 52.11 |  | 166 | 771 | 608 | 188 | 100 | 84 | 33 | 25 | 21 | 8 | 6 | 10 | 7 |
| 55.03 |  | 493 | 1286 | 622 | 189 | 66 | 36 | 11 | 14 | 5 | 3 | 1 | 3 | 0 |
| 56.05 |  | 1509 | 1174 | 435 | 124 | 20 | 16 | 14 | 6 | 2 | 9 | 3 | 1 | 1 |
| 52.06 |  | 621 | 1445 | 710 | 307 | 174 | 38 | 16 | 11 | 11 | 6 | 17 | 1 | 1 |
| 43.24 |  | 0 | 1292 | 1704 | 570 | 163 | 56 | 27 | 15 | 1 | 1 | 1 | 4 | 0.6 |
| 42.81 |  | 16 | 538 | 929 | 1273 | 315 | 160 | 50 | 19 | 12 | 2 | 7 | 1 | 3 |
| UK-CBT |  | UK(E+W) VIIf Beam trawl |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1991 | 2009 |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 10 |  | 1 |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 40.81 |  | 0 | 52 | 98 | 189 | 171 | 60 | 67 | 23 | 20 | 16 | 13 | 5 | 4 |
| 35.78 |  | 0 | 18 | 220 | 103 | 83 | 69 | 22 | 21 | 10 | 13 | 5 | 3 | 1 |
| 39.64 |  | 1.9 | 6 | 83 | 198 | 77 | 50 | 41 | 11 | 24 | 9 | 5 | 4 | 3 |
| 37.03 |  | 0 | 23 | 80 | 59 | 116 | 36 | 31 | 19 | 11 | 15 | 8 | 5 | 5 |
| 37.59 |  | 0 | 16 | 87 | 73 | 56 | 105 | 24 | 30 | 23 | 8 | 8 | 4 | 5 |
| 39.78 |  | 0.2 | 22 | 96 | 128 | 70 | 45 | 53 | 15 | 13 | 12 | 4 | 9 | 5 |
| 43 |  | 0 | 10 | 60 | 86 | 69 | 53 | 27 | 39 | 11 | 11 | 5 | 5 | 3 |
| 47.84 |  | 0 | 13 | 101 | 73 | 77 | 50 | 17 | 13 | 20 | 7 | 6 | 4 | 2 |
| 50.87 |  | 0.4 | 31 | 204 | 107 | 52 | 50 | 28 | 13 | 6 | 10 | 4 | 2 | 1 |
| 51.19 |  | 0.1 | 72 | 152 | 150 | 75 | 27 | 28 | 20 | 9 | 4 | 8 | 3 | 2 |
| 49.32 |  | 0 | 37 | 272 | 99 | 89 | 48 | 19 | 17 | 11 | 9 | 3 | 7 | 1 |
| 37.53 |  | 0 | 11 | 149 | 375 | 90 | 63 | 28 | 18 | 14 | 9 | 6 | 4 | 4 |
| 40.71 |  | 0.1 | 18 | 101 | 176 | 369 | 77 | 45 | 18 | 6 | 7 | 3 | 4 | 1 |
| 32.37 |  | 0 | 19 | 91 | 65 | 114 | 180 | 34 | 27 | 15 | 7 | 3 | 5 | 1 |
| 27.73 |  | 0 | 27 | 78 | 126 | 55 | 60 | 115 | 15 | 14 | 4 | 5 | 2 | 2 |
| 18.57 |  | 0 | 16 | 86 | 94 | 103 | 32 | 39 | 69 | 13 | 8 | 4 | 2 | 2 |
| 15.37 |  | 0.9 | 18 | 77 | 89 | 77 | 82 | 32 | 41 | 76 | 8 | 8 | 4 | 2 |
| 13.83 |  | 0 | 12 | 76 | 100 | 67 | 52 | 54 | 19 | 32 | 42 | 10 | 5 | 2 |
| 12.23 |  | 0 | 22 | 54 | 73 | 73 | 63 | 28 | 29 | 12 | 12 | 29 | 4 | 3 |

Table 7.13.9-Sole in VIIfg. Tuning series - continued
Indices in bold are used in the assessment

| UK-(BTS-3Q) | UK(E+W) VIIf Corystes (automated indices since 1995) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 2009 |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.85 |  |  |  |  |  |  |  |
| 0 | 9 |  |  |  |  |  |  |  |  |  |
| 74.120 | 22 | 60 | 242 | 36 | 14 | 4 | 0 | 0 | 0 | 0 |
| 91.909 | 132 | 204 | 304 | 162 | 18 | 14 | 6 | 4 | 2 | 2 |
| 69.858 | 21 | 269 | 219 | 35 | 11 | 3 | 5 | 2 | 0 | 0 |
| 123.410 | 40 | 297 | 638 | 83 | 21 | 18 | 5 | 0 | 3 | 2 |
| 125.078 | 5 | 493 | 325 | 174 | 37 | 23 | 12 | 1 | 2 | 1 |
| 127.672 | 6 | 207 | 436 | 52 | 28 | 3 | 2 | 2 | 1 | 1 |
| 120.816 | 1 | 424 | 430 | 133 | 23 | 11 | 9 | 0 | 0 | 3 |
| 114.886 | 31 | 142 | 255 | 60 | 13 | 7 | 14 | 1 | 1 | 1 |
| 118.592 | 3 | 178 | 251 | 64 | 27 | 7 | 3 | 4 | 1 | 3 |
| 114.886 | 37 | 498 | 207 | 21 | 13 | 14 | 5 | 3 | 6 | 0 |
| 114.886 | 104 | 885 | 472 | 57 | 11 | 9 | 5 | 2 | 1 | 5 |
| 118.592 | 29 | 2922 | 297 | 38 | 16 | 7 | 4 | 5 | 1 | 0 |
| 118.592 | 16 | 1086 | 1608 | 37 | 26 | 6 | 0 | 2 | 1 | 1 |
| 118.592 | 26 | 449 | 711 | 307 | 23 | 9 | 6 | 2 | 0 | 2 |
| 118.592 | 9 | 786 | 283 | 151 | 121 | 14 | 7 | 2 | 3 | 0 |
| 118.592 | 14 | 465 | 628 | 55 | 30 | 56 | 9 | 3 | 3 | 0 |
| 114.886 | 63 | 862 | 434 | 99 | 15 | 22 | 42 | 4 | 3 | 0 |
| 118.592 | 44 | 407 | 267 | 38 | 16 | 7 | 5 | 17 | 1 | 2 |
| 118.592 | 13 | 324 | 238 | 47 | 16 | 8 | 0 | 2 | 12 | 0 |
| 118.592 | 104 | 424 | 128 | 51 | 16 | 13 | 7 | 3 | 4 | 14 |
| 118.592 | 6 | 1232 | 124 | 15 | 18 | 7 | 9 | 4 | 3 | 5 |
| 118.592 | 1 | 604 | 377 | 29 | 8 | 10 | 4 | 3 | 3 | 2 |

IR - GFS : Irish Groundfish Survey (IBTS 4th Qtr) - VIIb Sole number at age (Interim indices for new Celtic Explorer series) 20032009

|  | 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 10 |  |  |  |  |  |  |  |  |  |
| 832 |  | 1.0 | 5.2 | 1.1 | 3.2 | 3.0 | 4.1 | 4.0 | 0.0 | 1.0 | 0.0 |
| 980 |  | 1.0 | 8.0 | 6.0 | 5.0 | 1.0 | 2.0 | 1.0 | 0.0 | 0.0 | 1.0 |
| 845 |  | 0.0 | 0.0 | 6.0 | 2.0 | 4.0 | 2.0 | 2.0 | 0.0 | 0.0 | 0.0 |
| 1046 |  | 0.0 | 0.0 | 4.0 | 4.0 | 6.0 | 4.0 | 1.0 | 0.0 | 0.0 | 0.0 |
| 1168 |  | 0.0 | 2.0 | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 | 0.0 | 0.0 | 0.0 |
| 1139 |  | 2.0 | 9.0 | 7.0 | 3.0 | 2.0 | 0.0 | 2.0 | 0.0 | 1.0 | 0.0 |
| 1018 |  | 0.0 | 15.0 | 3.0 | 4.0 | 1.0 | 1.0 | 2.0 | 1.0 | 0.0 | 2.0 |

UK (E+W) TRAWL 107F. (Processed as unsexed - from 2001WG)

| 1991 | 2009 |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |
| 1 | 10 |  |  |  |  |  |  |  |  |  |
| 18.57 | 0 | 1.7 | 6.4 | 13 | 11.2 | 3.5 | 3.3 | 1.1 | 0.8 | 0.8 |
| 16 | 0 | 8.4 | 29.4 | 10.4 | 6.9 | 5.9 | 1.5 | 1.8 | 0.8 | 0.9 |
| 13.79 | 0.1 | 0.8 | 3.7 | 10.2 | 3.8 | 2 | 1.4 | 0.3 | 0.6 | 0.2 |
| 9.48 | 0 | 1.7 | 4.3 | 2.5 | 4.9 | 1.7 | 1.5 | 1.1 | 0.6 | 0.7 |
| 8.46 | 0 | 2.3 | 12 | 5.3 | 2.5 | 4.5 | 0.9 | 1.2 | 0.7 | 0.2 |
| 8.67 | 0.1 | 2.8 | 4.3 | 4.9 | 2.4 | 1.4 | 1.4 | 0.3 | 0.5 | 0.2 |
| 8.14 | 0 | 2 | 8 | 6.8 | 4.1 | 2.1 | 0.7 | 1.2 | 0.4 | 0.3 |
| 7.13 | 0 | 2 | 4 | 2.7 | 2.1 | 1.3 | 0.4 | 0.3 | 0.5 | 0.1 |
| 5.69 | 0.1 | 8.5 | 12.4 | 3.5 | 1.5 | 1.2 | 0.8 | 0.4 | 0.1 | 0.3 |
| 4.05 | 0 | 0.9 | 1.8 | 1.6 | 0.7 | 0.2 | 0.2 | 0.2 | 0.1 | 0 |
| 4.42 | 0 | 1.5 | 10.1 | 2.3 | 1.7 | 0.6 | 0.3 | 0.2 | 0.2 | 0.1 |
| 6.1 | 0 | 0.5 | 4.8 | 8.2 | 1.8 | 1 | 0.3 | 0.2 | 0.2 | 0.1 |
| 9.94 | 0.1 | 1.6 | 2.8 | 3.3 | 6.7 | 1 | 0.7 | 0.3 | 0.1 | 0.1 |
| 9.42 | 0 | 1 | 4.8 | 2.9 | 3.3 | 4.9 | 0.9 | 0.6 | 0.4 | 0.2 |
| 12.09 | 0 | 2.6 | 4.9 | 6.1 | 2.3 | 2.6 | 4.9 | 0.7 | 0.7 | 0.2 |
| 12.97 | 0 | 0.4 | 7.1 | 7.7 | 9.5 | 3 | 3.9 | 6.9 | 1.3 | 0.9 |
| 10.66 | 0 | 0.5 | 2.6 | 3.5 | 3.2 | 3.2 | 1.2 | 1.5 | 2.6 | 0.3 |
| 10.13 | 0 | 0.4 | 3.5 | 5 | 3.8 | 2.9 | 2.7 | 0.9 | 1.6 | 2.2 |
| 8.97 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 7.13.10 - Sole VIIfg - XSA diagnostics
Lowestoft VPA Version 3.1
21/04/2010 17:39
Extended Survivors Analysis
CELTIC SEA SOLE 2010WG COMBSEX PLUSGROUP
CPUE data from file SOL7FTUN.txt
Catch data for 39 years. 1971 to 2009. Ages 1 to 10 .


Time series weights :
Tapered time weighting not applied

Catchability analysis
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=7$

Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.500$

Minimum standard error for population
estimates derived from each fleet $=.300$
Prior weighting not applied

Tuning converged after 48 iterations
1

Regression weights

| Fishing mortalities | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |  |  |  | 2009 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.142 | 0.11 | 0.008 | 0.02 | 0.096 | 0.051 | 0.136 | 0.09 | 0.057 | 0.072 |
| 3 | 0.416 | 0.208 | 0.311 | 0.243 | 0.396 | 0.253 | 0.334 | 0.227 | 0.15 | 0.153 |
| 4 | 0.383 | 0.402 | 0.346 | 0.374 | 0.425 | 0.303 | 0.309 | 0.281 | 0.198 | 0.185 |
| 5 | 0.318 | 0.4 | 0.544 | 0.488 | 0.432 | 0.43 | 0.313 | 0.308 | 0.314 | 0.163 |
| 6 | 0.233 | 0.538 | 0.372 | 0.667 | 0.314 | 0.355 | 0.229 | 0.26 | 0.262 | 0.237 |
| 7 | 0.345 | 0.376 | 0.37 | 0.59 | 0.353 | 0.219 | 0.193 | 0.319 | 0.221 | 0.232 |
| 8 | 0.478 | 0.381 | 0.543 | 0.538 | 0.273 | 0.219 | 0.149 | 0.262 | 0.222 | 0.153 |
| 9 | 0.618 | 0.527 | 0.641 | 0.484 | 0.408 | 0.271 | 0.286 | 0.203 | 0.283 | 0.171 |

XSA population numbers (Thousands)


Table 7.13.10 - Sole VIIfg - XSA diagnostics - continued
Log catchability residuals

Fleet : BEL-CBT
Age

|  | 1971 | 1972 | 1973 |
| :--- | :---: | :---: | ---: |
| 1 | No data for this fleet at this age |  |  |
| 2 | 0.23 | 0.14 | 0.54 |
| 3 | -0.48 | 0.18 | 0.38 |
| 4 | 0.26 | -0.16 | 0.13 |
| 5 | 0.32 | 0.14 | 0.2 |
| 6 | 0.13 | 0.3 | -0.09 |
| 7 | 0.5 | -0.01 | -0.3 |
| 8 | 0.32 | 0.21 | -0.42 |
| 9 | 0.02 | -0.1 | -0.18 |


| 1974 | 1975 |
| ---: | ---: |
|  |  |
| 0.11 | -0.15 |
| -0.1 | -0.34 |
| -0.05 | -0.31 |
| 0.14 | 0 |
| 0.51 | 0.27 |
| 0.12 | 0.38 |
| -0.01 | -0.45 |
| 0.15 | -0.1 |

1976

0.55
0.4
-0.01
0.26
-0.18
0.15
0.57
0.07

| 1977 | 1978 | 1979 |
| ---: | ---: | ---: |
|  |  |  |
| 0.21 | 0.38 | 0.41 |
| 0.15 | 0.08 | 0.08 |
| -0.02 | 0.07 | 0.41 |
| -0.08 | -0.46 | 0.13 |
| 0.08 | -0.21 | 0.05 |
| 0.19 | -0.38 | 0.63 |
| -0.01 | -0.17 | 0.3 |
| -0.27 | -0.23 | 0.02 |

Age

|  | 1980 | 1981 | 1982 |
| :--- | :---: | :---: | ---: |
| 1 | No data for this fleet at this age |  |  |
| 2 | 1.18 | 0.55 | 0.22 |
| 3 | 0.05 | 0.22 | 0.12 |
| 4 | 0.27 | -0.09 | -0.15 |
| 5 | 0.21 | -0.13 | 0.05 |
| 6 | -0.04 | 0.21 | 0.21 |
| 7 | -0.87 | 0.17 | 0.41 |
| 8 | -0.16 | -0.14 | 0.36 |
| 9 | -0.01 | 0.08 | 0.42 |


| 1983 | 1984 | 1985 |
| ---: | ---: | ---: |
|  |  |  |
| 0.45 | 0.17 | -1.66 |
| -0.02 | -0.19 | -0.05 |
| -0.25 | -0.34 | -0.12 |
| -0.24 | 0.02 | 0.12 |
| -0.18 | -0.1 | 0.07 |
| 0.14 | 0.22 | -0.06 |
| 0.5 | -0.08 | 0.19 |
| -0.22 | -0.29 | -0.06 |


| 1986 | 1987 | 1988 |
| ---: | ---: | ---: |
|  |  |  |
| -0.09 | 0.42 | 0.05 |
| 0.01 | -0.16 | -0.54 |
| -0.09 | 0 | -0.19 |
| -0.04 | 0 | -0.05 |
| 0.11 | 0.38 | -0.02 |
| 0.05 | 0.69 | 0.02 |
| -0.27 | -0.13 | 0.57 |
| -0.08 | 0.16 | 0.03 |


|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :--- | :---: | ---: | ---: | ---: | ---: | ---: |
| 1 | No data for this fleet at this age |  |  |  |  |  |
| 2 | 0.09 | 1.61 | 0.79 | 0.42 | -0.16 | -1.11 |
| 3 | 0.18 | 0.42 | 0.42 | 0.29 | -0.2 | 0.1 |
| 4 | 0.13 | 0.08 | 0.31 | -0.03 | 0.23 | 0.42 |
| 5 | -0.04 | 0 | 0.24 | -0.18 | 0.19 | 0.05 |
| 6 | 0.22 | -0.35 | 0.02 | -0.34 | 0.36 | -0.03 |
| 7 | 0.2 | -0.45 | -0.85 | 0.23 | -0.08 | 0.1 |
| 8 | 0.25 | -0.41 | -0.97 | 0.44 | -0.74 | -0.02 |
| 9 | -0.16 | -0.4 | -0.47 | 0.29 | -0.01 | -0.29 |


| 1996 | 1997 | 1998 |
| ---: | ---: | ---: |
|  |  |  |
| -0.77 | -0.44 | -0.91 |
| 0.25 | 0.07 | 0 |
| 0.19 | -0.08 | 0.45 |
| 0.04 | 0.02 | -0.07 |
| 0.03 | 0.21 | -0.09 |
| -0.32 | 0.21 | 0.66 |
| -0.27 | -0.25 | 0.16 |
| -0.32 | 0.07 | -0.4 |

Age

|  | 2000 | 2001 | 2002 |
| :--- | :---: | :---: | ---: |
| 1 | No data for this fleet at this age |  |  |
| 2 | 0.26 | 0.09 | 99.99 |
| 3 | -0.03 | -0.7 | 0.04 |
| 4 | -0.55 | -0.17 | -0.2 |
| 5 | -0.92 | -0.31 | 0.39 |
| 6 | -1.6 | 0.07 | -0.2 |
| 7 | -1.28 | -0.4 | -0.22 |
| 8 | -0.82 | -0.74 | -0.01 |
| 9 | -0.6 | -0.39 | -0.02 |

2003

-3.29
-0.33
-0.06
0.06
0.57
0.45
0.22
0.27

| 2004 | 2005 |
| :--- | ---: |
|  |  |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |
| 99.99 | 99.99 |


| 2006 | 2007 | 2008 |
| ---: | ---: | ---: |
|  |  |  |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |
| 99.99 | 99.99 | 99.99 |


99.99
99.99

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -6.3729 | -5.1053 | -4.8863 | -4.9165 | -4.9817 | -5.0695 | -5.0695 | -5.0695 |
| S.E(Log q) | 0.866 | 0.2855 | 0.2321 | 0.2419 | 0.3734 | 0.4505 | 0.4193 | 0.2561 |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age Slope t-value Intercept RSquare No Pts Regs.e Mean Q

| 2 | 0.98 | 0.035 | 6.4 | 0.15 | 32 | 0.87 | -6.37 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 1.06 | -0.415 | 4.91 | 0.58 | 33 | 0.31 | -5.11 |
| 4 | 1.07 | -0.599 | 4.69 | 0.71 | 33 | 0.25 | -4.89 |
| 5 | 0.85 | 1.94 | 5.28 | 0.84 | 33 | 0.2 | -4.92 |
| 6 | 0.76 | 2.371 | 5.39 | 0.76 | 33 | 0.27 | -4.98 |
| 7 | 0.81 | 1.746 | 5.28 | 0.74 | 33 | 0.36 | -5.07 |
| 8 | 0.89 | 1.302 | 5.21 | 0.83 | 33 | 0.37 | -5.14 |
| 9 | 0.92 | 2.089 | 5.19 | 0.96 | 33 | 0.2 | -5.17 |
| 1 |  |  |  |  |  |  |  |

Table 7.13.10 - Sole VIIfg - XSA diagnostics - continued
Fleet: UK-CBT

| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 99.99 | 0.38 | 0.13 | -1.15 | 0.26 | 0.13 | 0.42 | -0.64 | -0.79 | -0.09 |
|  | 3 | 99.99 | 0.07 | 0.33 | -0.13 | -0.22 | -0.08 | 0.22 | -0.32 | -0.12 | 0.25 |
|  | 4 | 99.99 | 0.56 | 0.15 | 0.05 | -0.44 | -0.31 | 0.33 | 0.11 | -0.11 | -0.11 |
|  | 5 | 99.99 | 0.57 | 0.09 | -0.04 | -0.18 | -0.2 | 0.03 | 0.07 | 0.24 | -0.05 |
|  | 6 | 99.99 | 0.4 | 0.17 | -0.22 | -0.36 | 0.19 | -0.01 | 0.25 | 0.15 | 0.17 |
|  | 7 | 99.99 | 0.37 | -0.04 | 0.08 | -0.18 | -0.14 | 0.03 | 0.09 | -0.15 | 0.07 |
|  | 8 | 99.99 | 0.41 | -0.21 | -0.34 | -0.06 | 0.44 | -0.05 | 0.28 | 0.01 | 0.3 |
|  | 9 | 99.99 | 0.53 | 0.24 | 0.34 | 0.4 | 0.74 | 0.28 | 0.22 | 0.14 | -0.17 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 t at this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 | -0.12 | -0.11 | -0.47 | -0.55 | 0.01 | 0.34 | 0.27 | 0.88 | 0.59 | 0.51 |
|  | 3 | -0.16 | -0.5 | -0.15 | -0.12 | -0.42 | -0.17 | 0.17 | 0.29 | 0.63 | 0.41 |
|  | 4 | -0.05 | -0.62 | -0.1 | -0.2 | -0.48 | -0.06 | 0.23 | 0.23 | 0.4 | 0.41 |
|  | 5 | -0.22 | -0.38 | -0.2 | -0.01 | -0.24 | -0.29 | 0.22 | 0.3 | 0.12 | 0.16 |
|  | 6 | -0.37 | -0.32 | -0.17 | 0.05 | -0.21 | -0.43 | -0.21 | 0.36 | 0.19 | 0.35 |
|  | 7 | 0.07 | -0.35 | -0.08 | -0.08 | 0.03 | -0.19 | -0.14 | 0.3 | 0.33 | -0.02 |
|  | 8 | 0.27 | 0.07 | 0.42 | -0.06 | 0.19 | -0.24 | -0.01 | 0.42 | 0.25 | 0.12 |
|  | 9 | 0.59 | 0.36 | 0.74 | -0.14 | 0.57 | 0.06 | 0.37 | 0.55 | 0.65 | 0.21 |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean $\log$ q | -8.9437 | -6.9296 | -6.359 | -6.031 | -5.8216 | -5.7627 | -5.7627 | -5.7627 |
| S.E(Log q) | 0.5221 | 0.2989 | 0.3229 | 0.2386 | 0.274 | 0.1865 | 0.2703 | 0.4482 |

Regression statistics

Ages with q independent of year class strength and constant w.r.t. time.

|  | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.95 | 0.167 | 8.92 | 0.38 | 19 | 0.51 | -8.94 |
| 3 | 1.3 | -1.151 | 6.52 | 0.47 | 19 | 0.38 | -6.93 |
| 4 | 1.04 | -0.194 | 6.3 | 0.58 | 19 | 0.35 | -6.36 |
| 5 | 0.99 | 0.059 | 6.04 | 0.82 | 19 | 0.24 | -6.03 |
| 6 | 1 | 0.01 | 5.82 | 0.81 | 19 | 0.28 | -5.82 |
| 7 | 0.96 | 0.643 | 5.78 | 0.93 | 19 | 0.18 | -5.76 |
| 8 | 1.01 | -0.109 | 5.65 | 0.89 | 19 | 0.25 | -5.65 |
| 9 | 0.92 | 0.99 | 5.37 | 0.91 | 19 | 0.25 | -5.41 |

Fleet: UK(BTS-3Q)

| Age |  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -1.35 | -0.15 |
|  |  |  | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.05 | 0.33 |
|  |  |  | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.35 | 1.12 |
|  |  |  | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.09 | 0.6 |
|  | 5 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.19 | 0.36 |
|  | 6 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.48 |
|  | 7 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.54 |
|  | 8 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.68 |
|  | 9 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 1.74 |
| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | 1 | -0.44 | -0.19 | 0.24 | -0.64 | 0.39 | -0.63 | -0.63 | 0.13 | 0.57 | 0.85 |
|  | 2 | 0.43 | 0.19 | 0.15 | 0.34 | 0.37 | 0.13 | 0.13 | -0.23 | 0.27 | -0.3 |
|  | 3 | 0.16 | 0.53 | 0.61 | -0.01 | 0.84 | 0.21 | 0.53 | -0.56 | 0.19 | -0.45 |
|  | 4 | -0.03 | 0.22 | 0.84 | -0.13 | 0.42 | -0.09 | 0.73 | 0.24 | 0.18 | 0.16 |
|  | 5 | -0.11 | 0.63 | 0.96 | -1.08 | -0.27 | 0.04 | 0.09 | 0.96 | 0.65 | 0.56 |
|  | 6 | 0.16 | 0.26 | 0.62 | -1.19 | 0.58 | 0.57 | -0.3 | 0.47 | 0.45 | 0.27 |
|  | 7 | 0.6 | 99.99 | -0.68 | -0.32 | 99.99 | -0.64 | 0.14 | 0.76 | 0.72 | 1.28 |
|  | 8 | 99.99 | 1.03 | -0.11 | -0.13 | 99.99 | -0.23 | -0.09 | 1.24 | 0.37 | 0.68 |
|  | 9 | 99.99 | 0.89 | 0.43 | -0.19 | 1.76 | 0.33 | 1.52 | 99.99 | 1.63 | 99.99 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 | 0.52 | 0.26 | 0.32 | 0.04 | 0.53 | -0.24 | -0.14 | 0.16 | 0.4 | 0 |
|  | 2 | 0.53 | 0.34 | -0.03 | 0.28 | 0.24 | -0.47 | -0.51 | -0.83 | -0.86 | -0.56 |
|  | 3 | -0.64 | 0.45 | 0.45 | -0.09 | 0.17 | -0.62 | -0.54 | -0.45 | -1.45 | -0.8 |
|  | 4 | 0.29 | -0.01 | 0.55 | -0.1 | -0.26 | -0.66 | -0.48 | -0.62 | -0.58 | -1.19 |
|  | 5 | -0.23 | -0.16 | 0.22 | 0.44 | 0.25 | -0.41 | -0.83 | -0.17 | -0.93 | -0.79 |
|  | 6 | 99.99 | 0.2 | -0.1 | 0.36 | 0.48 | -0.95 | 99.99 | -0.75 | -0.33 | -1.3 |
|  |  | 0.31 | 0.26 | -0.14 | -0.05 | 0.35 | 0.13 | -1.29 | -0.4 | -0.73 | -0.84 |
|  | 8 | 0.25 | 99.99 | 1.27 | 0.87 | 0.42 | -0.72 | 0.05 | -0.25 | -0.06 | -0.75 |
|  | 9 | 1.37 | 1.56 | 99.99 | 99.99 | 99.99 | 0.36 | 99.99 | 0.49 | 0.35 | -0.18 |

## Table 7.13.10 - Sole VIIfg - XSA diagnostics - continued

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -7.1906 | -7.2537 | -8.5374 | -9.1455 | -9.2541 | -9.0942 | -9.3451 | -9.3451 | -9.3451 |
| S.E(Log q) | 0.5135 | 0.4145 | 0.6218 | 0.4997 | 0.585 | 0.6257 | 0.6566 | 0.6647 | 1.1462 |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age |  | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.6 | 2.506 | 7.76 | 0.66 | 22 | 0.28 | -7.19 |
|  | 2 | 0.94 | 0.25 | 7.33 | 0.48 | 22 | 0.4 | -7.25 |
|  | 3 | 0.86 | 0.401 | 8.5 | 0.3 | 22 | 0.55 | -8.54 |
|  | 4 | 1.68 | -1.449 | 10.04 | 0.19 | 22 | 0.82 | -9.15 |
|  | 5 | 1.51 | -1.32 | 10.28 | 0.25 | 22 | 0.87 | -9.25 |
|  | 6 | 1.66 | -1.463 | 10.67 | 0.23 | 19 | 1.01 | -9.09 |
|  | 7 | 2.2 | -2.501 | 13.29 | 0.2 | 19 | 1.27 | -9.35 |
|  | 8 | 1.65 | -2.058 | 11.42 | 0.39 | 18 | 0.92 | -9.09 |
|  | 9 | 2.04 | -2.554 | 12.1 | 0.34 | 14 | 1.22 | -8.48 |
|  | 1 |  |  |  |  |  |  |  |

Terminal year survivor and $F$ summaries :
Age 1 Catchability constant w.r.t. time and dependent on age
Year class $=2008$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ |  | Var <br> Ratio |  | N |  | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL-CBT | 1 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| UK-CBT | 1 | 0 |  | 0 |  | 0 |  | 0 | 0 | 0 |
| UK(BTS-3Q) | 6624 | 0.525 |  | 0 |  | 0 |  | 1 | 1 | 0 |
| F shrinkage mean 0 |  | 1.5 |  |  |  |  |  |  | 0 | 0 |

Weighted prediction :

| Survivors at end of year | Int | Ext |  | N |  | Var |  | F |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | s.e | s.e |  |  |  | Ratio |  |  |  |
| 6624 | 0.53 |  | 0 |  | 1 |  | 0 |  | 0 |

Age 2 Catchability constant w.r.t. time and dependent on age
Year class $=2007$


## Table 7.13.10 - Sole VIIfg - XSA diagnostics - continued

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2006$

| Fleet | E | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled Weights |  | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL-CBT | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| UK-CBT | 4142 | 0.266 | 0.078 | 0.29 |  | 2 | 0.544 | 0.1 |
| UK(BTS-3Q) | 1543 | 0.293 | 0.327 | 1.12 |  | 3 | 0.436 | 0.249 |
| F shrinkage mean | 1390 | 1.5 |  |  |  |  | 0.02 | 0.273 |

Weighted prediction :

| Survivors at end of year | Int | Ext | N | Var |  | F |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | s.e |  |  | Ratio |  |  |
|  | 0.2 | 0.26 |  | 6 | 1.336 |  | 0.153 |

Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2005$


Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=2004$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var Ratio | N | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL-CBT | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK-CBT | 2413 | 0.174 | 0.056 | 0.32 | 4 | 0.674 | 0.127 |
| UK(BTS-3Q) | 1079 | 0.24 | 0.087 | 0.36 | 5 | 0.313 | 0.264 |
| F shrinkage mean | 752 | 1.5 |  |  |  | 0.013 | 0.36 |

Weighted prediction :


Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=2003$


Table 7.13.10 - Sole VIIfg - XSA diagnostics - continued
Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=2002$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N |  | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL-CBT | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| UK-CBT | 695 | 0.144 | 0.068 | 0.47 |  | 6 | 0.752 | 0.211 |
| UK(BTS-3Q) | 450 | 0.233 | 0.147 | 0.63 |  | 7 | 0.236 | 0.31 |
| F shrinkage mean | 547 | 1.5 |  |  |  |  | 0.012 | 0.262 |

Weighted prediction :

| Survivors |  | Int | Ext | $N$ |  | Var | F |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year |  | s.e | s.e |  |  | Ratio |  |
|  | 625 | 0.12 | 0.08 |  | 14 | 0.648 | 0.232 |

Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 7
Year class $=2001$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N |  | Scaled Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BEL-CBT | 22 | 0.879 | 0 | 0 |  | 1 | 0.006 | 1.696 |
| UK-CBT | 692 | 0.135 | 0.092 | 0.68 |  | 7 | 0.771 | 0.131 |
| UK(BTS-3Q) | 356 | 0.234 | 0.166 | 0.71 |  | 8 | 0.213 | 0.241 |
| F shrinkage mean | 384 | 1.5 |  |  |  |  | 0.01 | 0.225 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |  |
| 585 | 0.12 | 0.12 | 17 | 1.003 |  |  |  |  |

Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 7
Year class $=2000$


Table 7.13.11 - Sole in VIIfg. Fishing mortality
Run title : CELTIC SEA SOLE,2010WG,COMBSEX,PLUSGROUP At 21/04/2010 17:41


Table 7.13.12-Sole in VIlfg. Stock numbers at age (start of year, in thousand)
Run title : CELTIC SEA SOLE,2010WG,COMBSEX,PLUSGROUP
At 21/04/2010 17:41

|  |  |  |  |
| :---: | :---: | :---: | :---: |
|  |  |  |  <br>  |
|  |  |  | $\stackrel{\omega}{\vec{A}}$ 성 |
| N |  |  <br>  |  |
|  |  | N |  |
|  | $\stackrel{\rightharpoonup}{\stackrel{\rightharpoonup}{*}} \underset{\omega}{\omega} \text { O }$ |  |  |
|  | Tr |  |  |
|  |  | $\stackrel{\rightharpoonup}{0} \underset{\omega}{0} \underset{\omega}{\infty} \underset{\sim}{\sim} \underset{\omega}{\sim} \underset{\sim}{\sim}$ |  |
|  |  |  |  |
|  |  | © | N్ర్ઠિ |
|  |  |  |  |
|  <br> $N \omega c \infty$ <br>  |  |  |  |

Table 7.13.13 - Sole in VIIfg. Summary
Run title : CELTIC SEA SOLE,2010WG,COMBSEX,PLUSGROUP At 21/04/2010 17:41

|  | RECRUITS Age 1 | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 4-8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 9606 | 9497 | 8030 | 1861 | 0.232 | 0.362 |
| 1972 | 4274 | 7990 | 6333 | 1278 | 0.202 | 0.265 |
| 1973 | 3386 | 6632 | 5298 | 1391 | 0.263 | 0.231 |
| 1974 | 3403 | 6696 | 5676 | 1105 | 0.195 | 0.233 |
| 1975 | 2972 | 5883 | 5028 | 919 | 0.183 | 0.198 |
| 1976 | 5192 | 5386 | 4359 | 1350 | 0.310 | 0.362 |
| 1977 | 4635 | 5938 | 4675 | 961 | 0.206 | 0.245 |
| 1978 | 5492 | 5082 | 3762 | 780 | 0.207 | 0.185 |
| 1979 | 3534 | 5094 | 3884 | 954 | 0.246 | 0.265 |
| 1980 | 5131 | 5243 | 4021 | 1314 | 0.327 | 0.291 |
| 1981 | 4858 | 4597 | 3421 | 1212 | 0.354 | 0.345 |
| 1982 | 4888 | 4807 | 3557 | 1128 | 0.317 | 0.332 |
| 1983 | 6789 | 5135 | 3657 | 1373 | 0.375 | 0.438 |
| 1984 | 4704 | 5374 | 3916 | 1266 | 0.323 | 0.396 |
| 1985 | 5656 | 4790 | 3307 | 1328 | 0.402 | 0.422 |
| 1986 | 3157 | 4622 | 3367 | 1600 | 0.475 | 0.525 |
| 1987 | 5738 | 3733 | 2517 | 1222 | 0.486 | 0.551 |
| 1988 | 4491 | 3903 | 2708 | 1146 | 0.423 | 0.536 |
| 1989 | 3719 | 3245 | 2109 | 992 | 0.470 | 0.511 |
| 1990 | 8607 | 3882 | 2403 | 1189 | 0.495 | 0.629 |
| 1991 | 4199 | 3601 | 2128 | 1107 | 0.520 | 0.458 |
| 1992 | 4455 | 3858 | 2447 | 981 | 0.401 | 0.387 |
| 1993 | 4426 | 3836 | 2478 | 928 | 0.375 | 0.438 |
| 1994 | 3410 | 3265 | 2257 | 1009 | 0.447 | 0.507 |
| 1995 | 3319 | 3086 | 2155 | 1157 | 0.537 | 0.633 |
| 1996 | 4051 | 3061 | 2081 | 995 | 0.478 | 0.562 |
| 1997 | 5478 | 2975 | 1821 | 927 | 0.509 | 0.658 |
| 1998 | 6293 | 3058 | 1625 | 875 | 0.539 | 0.656 |
| 1999 | 15138 | 4282 | 1821 | 1012 | 0.556 | 0.556 |
| 2000 | 7860 | 3895 | 1942 | 1091 | 0.562 | 0.351 |
| 2001 | 4177 | 5409 | 3121 | 1168 | 0.374 | 0.419 |
| 2002 | 6896 | 5970 | 4095 | 1345 | 0.328 | 0.435 |
| 2003 | 5413 | 5640 | 3772 | 1392 | 0.369 | 0.531 |
| 2004 | 6325 | 5219 | 3539 | 1249 | 0.353 | 0.359 |
| 2005 | 6273 | 5471 | 3581 | 1044 | 0.292 | 0.305 |
| 2006 | 4502 | 4895 | 3190 | 946 | 0.297 | 0.239 |
| 2007 | 4389 | 4846 | 3503 | 945 | 0.270 | 0.286 |
| 2008 | 9997 | 5331 | 3436 | 800 | 0.2328 | 0.2432 |
| 2009 | 7321 | 6475 | 4180 | 790 | 0.189 | 0.1941 |
| 2010 | $4998{ }^{1}$ | $6569^{2}$ | $4420^{2}$ |  |  | $0.1941^{3}$ |
| Arith. Mean 0 Units | 5491 (Thousands) | $\begin{array}{r} 4915 \\ \text { (Tonnes) } \end{array}$ | $\begin{array}{r} 3467 \\ \text { (Tonnes) } \end{array}$ | $\begin{array}{r} 1132 \\ \text { (Tonnes) } \end{array}$ | 0.362 | 0.3984 |

${ }^{1}$ Geometric mean 1971-2007
${ }^{3}$ From forecast
${ }^{4} \mathrm{~F}_{(07-09)}$ rescaled to $\mathrm{F}_{2009}$

Table 7.13.14 - Sole in VIIfg
Input for catch forecast and Fmsy analysis

| Input: | F mean 07-09 rescaled to F2009 |
| :--- | :--- |
|  | Catch and stock weights are mean 07-09 |
|  | Recruits age 1 in 2010,11 and $12 \mathrm{GM}(71-07)$ |


| Label | Value | CV | Label | Value | CV |
| :--- | ---: | ---: | :--- | ---: | ---: |
|  |  |  | Weight in the stock |  |  |
| Number at age |  |  | WS1 | 0.090 | 0.00 |
| N1 | 4997 | 0.34 | WS2 | 0.138 | 0.13 |
| N2 | 6624 | 0.53 | WS3 | 0.196 | 0.09 |
| N3 | 7614 | 0.29 | WS4 | 0.252 | 0.07 |
| N4 | 2633 | 0.26 | WS5 | 0.307 | 0.06 |
| N5 | 1972 | 0.29 | WS6 | 0.361 | 0.06 |
| N6 | 1847 | 0.14 | WS7 | 0.415 | 0.05 |
| N7 | 1027 | 0.15 | WS8 | 0.468 | 0.04 |
| N8 | 624 | 0.12 | WS9 | 0.519 | 0.03 |
| N9 | 585 | 0.12 | WS10 | 0.635 | 0.05 |


| H.cons selectivity |  |  |
| :--- | ---: | ---: |
| sH1 | 0.0000 | 0 |
| sH2 | 0.0590 | 0.22 |
| sH3 | 0.1420 | 0.14 |
| sH4 | 0.1780 | 0.10 |
| sH5 | 0.2110 | 0.21 |
| sH6 | 0.2040 | 0.15 |
| sH7 | 0.2070 | 0.14 |
| sH8 | 0.1710 | 0.08 |
| sH9 | 0.1760 | 0.25 |
| sH10 | 0.1760 | 0.25 |

Natural mortality

| M1 | 0.1 | 0.1 | MT1 | 0 | 0 |
| :--- | ---: | :--- | :--- | ---: | ---: |
| M2 | 0.1 | 0.1 | MT2 | 0.14 | 0.1 |
| M3 | 0.1 | 0.1 | MT3 | 0.45 | 0.1 |
| M4 | 0.1 | 0.1 | MT4 | 0.88 | 0.1 |
| M5 | 0.1 | 0.1 | MT5 | 0.98 | 0.1 |
| M6 | 0.1 | 0.1 | MT6 | 1 | 0 |
| M7 | 0.1 | 0.1 | MT7 | 1 | 0 |
| M8 | 0.1 | 0.1 | MT8 | 1 | 0 |
| M9 | 0.1 | 0.1 | MT9 | 1 | 0 |
| M10 | 0.1 | 0.1 | MT10 | 1 | 0 |
|  |  |  |  |  |  |
| Relative effort |  |  |  | Kear effect for natural mortality |  |
| in HC fihery |  |  |  |  |  |
| HF10 | 1 | 0.19 | K11 | 1 | 0.1 |
| HF11 | 1 | 0.19 | K12 | 1 | 0.1 |
| HF12 | 1 | 0.19 |  | 1 | 0.1 |

Recruitment in 2011 and 2012

| R11 | 4998 | 0.34 |
| :--- | :--- | :--- |
| R12 | 4998 | 0.34 |

Table 7.13.15 - Sole in VIIfg. Management option table
MFDP version 1a
Run: Sole VIlfg_Fin
CELTIC SEA SOLE 2010WG COMBSEX PLUSGROUP
Time and date: 18:22 13/05/2010
Fbar age range: 4-8

| 2010 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 6569 | 4420 | 1.0000 | 0.1941 | 866 |


| 2011 |  |  |  | 2012 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 6842 | 5053 | 0.0000 | 0.0000 | 0 | 7998 | 6382 |
| . | 5053 | 0.1000 | 0.0194 | 102 | 7885 | 6274 |
| . | 5053 | 0.2000 | 0.0388 | 203 | 7774 | 6169 |
| . | 5053 | 0.3000 | 0.0582 | 302 | 7665 | 6065 |
| . | 5053 | 0.4000 | 0.0777 | 399 | 7558 | 5963 |
| . | 5053 | 0.5000 | 0.0971 | 494 | 7453 | 5863 |
| . | 5053 | 0.6000 | 0.1165 | 588 | 7349 | 5765 |
| . | 5053 | 0.7000 | 0.1359 | 680 | 7248 | 5668 |
| . | 5053 | 0.8000 | 0.1553 | 770 | 7148 | 5573 |
| . | 5053 | 0.9000 | 0.1747 | 859 | 7050 | 5480 |
|  | 5053 | 1.0000 | 0.1941 | 947 | 6953 | 5388 |
|  | 5053 | 1.1000 | 0.2136 | 1033 | 6859 | 5299 |
| . | 5053 | 1.2000 | 0.2330 | 1118 | 6765 | 5210 |
|  | 5053 | 1.3000 | 0.2524 | 1201 | 6674 | 5124 |
|  | 5053 | 1.4000 | 0.2718 | 1282 | 6584 | 5038 |
| . | 5053 | 1.5000 | 0.2912 | 1363 | 6496 | 4955 |
| . | 5053 | 1.6000 | 0.3106 | 1441 | 6409 | 4872 |
|  | 5053 | 1.7000 | 0.3300 | 1519 | 6324 | 4792 |
| . | 5053 | 1.8000 | 0.3495 | 1595 | 6240 | 4712 |
|  | 5053 | 1.9000 | 0.3689 | 1670 | 6158 | 4634 |
|  | 5053 | 2.0000 | 0.3883 | 1744 | 6077 | 4558 |

Input units are thousands and kg - output in tonnes

| Fmult corresponding to $\mathrm{Fpa}=1.91$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 5053 | 1.91 | 0.37 | 1675 | 6153 | 4630 |
| Fmult corresponding to Fmsy $=1.6$ |  |  |  |  |  |  |
|  | 5053 | 1.6 | 0.3106 | 1441 | 6409 | 4872 |
| $\mathrm{Bpa}=2200 \mathrm{t}$ |  |  |  |  |  |  |

Table 7.13.16-Sole in VIIfg. Detailed results
MFDP version 1a
Run: Sole VIIfg_Fin
Time and date: 18:22 13/05/2010
Fbar age range: 4-8

| Year: Age | 2010 F | F multiplier: CatchNos | 1 | Yield | Fbar: StockNos | $\begin{aligned} & 0.194 \\ & \quad \text { Biomass } \\ & \hline \end{aligned}$ | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.000 | 0 |  | 0 | 4998 | 450 | 0 | 0 | 0 | 0 |
| 2 | 0.059 | 362 |  | 57 | 6624 | 916 | 927 | 128 | 927 | 128 |
| 3 | 0.142 | 962 |  | 203 | 7615 | 1490 | 3427 | 671 | 3427 | 671 |
| 4 | 0.178 | 410 |  | 108 | 2634 | 664 | 2318 | 584 | 2318 | 584 |
| 5 | 0.211 | 357 |  | 112 | 1972 | 605 | 1933 | 593 | 1933 | 593 |
| 6 | 0.204 | 325 |  | 117 | 1847 | 667 | 1847 | 667 | 1847 | 667 |
| 7 | 0.208 | 183 |  | 75 | 1027 | 426 | 1027 | 426 | 1027 | 426 |
| 8 | 0.171 | 94 |  | 42 | 625 | 292 | 625 | 292 | 625 | 292 |
| 9 | 0.176 | 90 |  | 45 | 585 | 304 | 585 | 304 | 585 | 304 |
| 10 | 0.176 | 183 |  | 107 | 1187 | 754 | 1187 | 754 | 1187 | 754 |
| Total |  | 2965 |  | 866 | 29114 | 6569 | 13876 | 4420 | 13876 | 4420 |
| Year: Age | 2011 F | F multiplier: CatchNos | 1 | Yield | Fbar: StockNos | $0.194$ <br> Biomass | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.000 | 0 |  | 0 | 4998 | 450 | 0 | 0 | 0 | 0 |
| 2 | 0.059 | 247 |  | 39 | 4522 | 626 | 633 | 88 | 633 | 88 |
| 3 | 0.142 | 714 |  | 151 | 5650 | 1105 | 2542 | 497 | 2542 | 497 |
| 4 | 0.178 | 929 |  | 244 | 5977 | 1506 | 5259 | 1325 | 5259 | 1325 |
| 5 | 0.211 | 361 |  | 113 | 1994 | 612 | 1955 | 600 | 1955 | 600 |
| 6 | 0.204 | 254 |  | 92 | 1446 | 522 | 1446 | 522 | 1446 | 522 |
| 7 | 0.208 | 243 |  | 99 | 1363 | 566 | 1363 | 566 | 1363 | 566 |
| 8 | 0.171 | 113 |  | 51 | 755 | 353 | 755 | 353 | 755 | 353 |
| 9 | 0.176 | 73 |  | 36 | 477 | 248 | 477 | 248 | 477 | 248 |
| 10 | 0.176 | 207 |  | 121 | 1344 | 854 | 1344 | 854 | 1344 | 854 |
| Total |  | 3142 |  | 947 | 28526 | 6842 | 15774 | 5053 | 15774 | 5053 |


| Year: Age | 2012 F | F multiplier: CatchNos | Yield | Fbar: StockNos | $\begin{aligned} & 0.194 \\ & \quad \text { Biomass } \\ & \hline \end{aligned}$ | SSNos(Jan) | SSB(Jan) | SSNos(ST) | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.000 | 0 | 0 | 4998 | 450 | 0 | 0 | 0 | 0 |
| 2 | 0.059 | 247 | 39 | 4522 | 626 | 633 | 88 | 633 | 88 |
| 3 | 0.142 | 487 | 103 | 3857 | 755 | 1736 | 340 | 1736 | 340 |
| 4 | 0.178 | 690 | 181 | 4434 | 1117 | 3902 | 983 | 3902 | 983 |
| 5 | 0.211 | 819 | 257 | 4525 | 1389 | 4435 | 1362 | 4435 | 1362 |
| 6 | 0.204 | 257 | 93 | 1462 | 528 | 1462 | 528 | 1462 | 528 |
| 7 | 0.208 | 191 | 78 | 1067 | 443 | 1067 | 443 | 1067 | 443 |
| 8 | 0.171 | 150 | 68 | 1002 | 469 | 1002 | 469 | 1002 | 469 |
| 9 | 0.176 | 89 | 44 | 576 | 299 | 576 | 299 | 576 | 299 |
| 10 | 0.176 | 213 | 125 | 1381 | 877 | 1381 | 877 | 1381 | 877 |
| Total |  | 3142 | 987 | 27826 | 6953 | 16195 | 5388 | 16195 | 5388 |

Input units are thousands and kg - output in tonnes


Sole VIlf,g : Year-class \% contribution to
a) 2011 landings

b) 2012 SSB


Table 7.13.18 - Sole in VIIfg. Yield per recruit summary table

MFYPR version 2a
Run: Sole Vlifg FinalYield
Time and date: 14:35 14/05/2010
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.000 | 0.000 | 0.000 | 10.508 | 4.394 | 8.178 | 4.082 | 8.178 | 4.082 |
| 0.1000 | 0.019 | 0.129 | 0.056 | 9.219 | 3.624 | 6.894 | 3.312 | 6.894 | 3.312 |
| 0.2000 | 0.039 | 0.224 | 0.094 | 8.270 | 3.066 | 5.949 | 2.756 | 5.949 | 2.756 |
| 0.3000 | 0.058 | 0.297 | 0.120 | 7.543 | 2.647 | 5.227 | 2.338 | 5.227 | 2.338 |
| 0.4000 | 0.078 | 0.355 | 0.139 | 6.969 | 2.322 | 4.657 | 2.014 | 4.657 | 2.014 |
| 0.5000 | 0.097 | 0.401 | 0.152 | 6.504 | 2.064 | 4.197 | 1.757 | 4.197 | 1.757 |
| 0.6000 | 0.117 | 0.440 | 0.162 | 6.120 | 1.856 | 3.818 | 1.550 | 3.818 | 1.550 |
| 0.7000 | 0.136 | 0.472 | 0.169 | 5.799 | 1.684 | 3.501 | 1.380 | 3.501 | 1.380 |
| 0.8000 | 0.155 | 0.499 | 0.175 | 5.525 | 1.542 | 3.232 | 1.238 | 3.232 | 1.238 |
| 0.9000 | 0.175 | 0.523 | 0.178 | 5.289 | 1.422 | 3.000 | 1.119 | 3.000 | 1.119 |
| 1.0000 | 0.194 | 0.544 | 0.181 | 5.084 | 1.320 | 2.800 | 1.018 | 2.800 | 1.018 |
| 1.1000 | 0.214 | 0.562 | 0.183 | 4.905 | 1.232 | 2.624 | 0.931 | 2.624 | 0.931 |
| 1.2000 | 0.233 | 0.578 | 0.184 | 4.746 | 1.156 | 2.470 | 0.856 | 2.470 | 0.856 |
| 1.3000 | 0.252 | 0.592 | 0.185 | 4.604 | 1.089 | 2.332 | 0.790 | 2.332 | 0.790 |
| 1.4000 | 0.272 | 0.605 | 0.186 | 4.477 | 1.031 | 2.210 | 0.733 | 2.210 | 0.733 |
| 1.5000 | 0.291 | 0.616 | 0.186 | 4.363 | 0.980 | 2.099 | 0.683 | 2.099 | 0.683 |
| 1.6000 | 0.311 | 0.627 | 0.186 | 4.259 | 0.934 | 1.999 | 0.638 | 1.999 | 0.638 |
| 1.7000 | 0.330 | 0.636 | 0.185 | 4.164 | 0.893 | 1.909 | 0.598 | 1.909 | 0.598 |
| 1.8000 | 0.350 | 0.645 | 0.185 | 4.078 | 0.857 | 1.826 | 0.562 | 1.826 | 0.562 |
| 1.9000 | 0.369 | 0.653 | 0.185 | 3.999 | 0.824 | 1.751 | 0.530 | 1.751 | 0.530 |
| 2.0000 | 0.388 | 0.661 | 0.184 | 3.926 | 0.794 | 1.682 | 0.501 | 1.682 | 0.501 |


| Reference point | F multiplier | Absolute $\mathbf{F}$ |
| :--- | :---: | :---: |
| Fbar(4-8) | 1.000 | 0.194 |
| FMax | 1.535 | 0.298 |
| F0.1 | 0.662 | 0.128 |
| F35\%SPR | 0.669 | 0.130 |
|  |  |  |
| Weights in kilograms |  |  |

Table 7.13.19 - Sole VIIfg - FMSY summary
Estimates of biomass and fishing mortality reference levels derived from the fit of three stock
and recruit relationships and the yield per recruit Fmsy proxies.

Stock name
SOLVIIFG.SUM
Sen filename
solviifg.sen
pf, pm $0 \quad 0$
Number of iterations 1000
Simulate variation in Biological parameters TRUE
SR relationship constrained TRUE

Ricker
936/1000 Iterations resulted in feasible parameter estimates

| Fcrash |  | Fmsy | Bmsy | MSY | ADMB Alph | ADMB Bet | Unscaled Alp | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis ${ }^{\text {P }}$ | 0.906 | 0.357 | 3019 | 1015 | 0.205 | 2.762 | 5.138 | 0.0003 | 32.557 |
| Mean | 0.846 | 0.322 | 3047 | 1004 | 0.220 | 2.686 | 5.071 | 0.0003 |  |
| 5\%ile | 0.512 | 0.222 | 2419 | 731 | 0.152 | 2.096 | 3.781 | 0.0003 |  |
| 25\%ile | 0.664 | 0.275 | 2730 | 878 | 0.187 | 2.444 | 4.456 | 0.0003 |  |
| 50\%ile | 0.796 | 0.314 | 2984 | 998 | 0.216 | 2.681 | 5.005 | 0.0003 |  |
| 75\%ile | 0.955 | 0.359 | 3284 | 1125 | 0.248 | 2.935 | 5.595 | 0.0004 |  |
| 95\%ile | 1.339 | 0.454 | 3911 | 1305 | 0.307 | 3.281 | 6.602 | 0.0004 |  |
| CV | 0.342 | 0.216 | 0.151 | 0.174 | 0.216 | 0.135 | 0.171 | 0.1351 |  |

Beverton-Holt
342/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alph | ADMB Beta | nscaled Alpl | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 1.018 | 0.176 | 7250 | 1160 | 1.656 | 1.927 | 7682 | 1313 | 42.969 |
| Mean | 2.588 | 0.198 | 10804 | 951 | 2.260 | 2.368 | 5689 | 404 |  |
| 5\%ile | 1.048 | 0.027 | 2062 | 648 | 1.869 | 2.067 | 4859 | 140 |  |
| 25\%ile | 1.731 | 0.159 | 3030 | 787 | 2.139 | 2.269 | 5328 | 219 |  |
| 50\%ile | 2.459 | 0.215 | 4235 | 916 | 2.272 | 2.368 | 5598 | 308 |  |
| 75\%ile | 3.368 | 0.257 | 6565 | 1080 | 2.387 | 2.477 | 5948 | 520 |  |
| 95\%ile | 4.478 | 0.323 | 55013 | 1346 | 2.618 | 2.692 | 6804 | 903 |  |
| CV | 0.423 | 0.437 | 1.615 | 0.250 | 0.099 | 0.079 | 0.110 | 0.680 |  |

Smooth hockeystick

|  | Fcrash | Fmsy | Bmsy |  | MSY |  | ADMB Alph | ADMB Beta Unsc | led Alpl | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 0.579 | 0.298 |  | 3374 |  | 938 | 0.982 | 0.469 | 1.556 | 1624 | 31.567 |
| Mean | 0.491 | 0.290 |  | 8656 |  | 965 | 0.922 | 0.506 | 1.460 | 1754 |  |
| 5\%ile | 0.342 | 0.029 |  | 1683 |  | 694 | 0.795 | 0.471 | 1.259 | 1633 |  |
| 25\%ile | 0.416 | 0.208 |  | 1997 |  | 828 | 0.873 | 0.481 | 1.383 | 1666 |  |
| 50\%ile | 0.478 | 0.315 |  | 2809 |  | 935 | 0.924 | 0.497 | 1.463 | 1724 |  |
| 75\%ile | 0.555 | 0.385 |  | 4870 |  | 1069 | 0.973 | 0.522 | 1.542 | 1811 |  |
| 95\%ile | 0.687 | 0.481 |  | 45548 |  | 1348 | 1.042 | 0.574 | 1.650 | 1989 |  |
| CV | 0.218 | 0.477 |  | 1.803 |  | 0.205 | 0.083 | 0.066 | 0.083 | 0.066 |  |
| Per recruit |  |  |  |  |  |  |  |  |  |  |  |
|  | F35 | F40 | F01 |  | Fmax |  | Bmsypr | MSYpr Fpa |  | Flim |  |
| Determinis 1 | 0.130 | 0.108 |  | 0.128 |  | 0.298 | 0.668 | 0.186 | 0.37 | 0.52 |  |
| Mean | 0.114 | 0.095 |  | 0.118 |  | 0.333 | 1.692 | 0.189 |  |  |  |
| 5\%ile | 0.002 | 0.001 |  | 0.002 |  | 0.029 | 0.326 | 0.136 |  |  |  |
| 25\%ile | 0.055 | 0.045 |  | 0.060 |  | 0.207 | 0.390 | 0.164 |  |  |  |
| 50\%ile | 0.123 | 0.103 |  | 0.132 |  | 0.319 | 0.553 | 0.185 |  |  |  |
| 75\%ile | 0.169 | 0.142 |  | 0.171 |  | 0.411 | 0.949 | 0.208 |  |  |  |
| 95\%ile | 0.218 | 0.183 |  | 0.221 |  | 0.684 | 8.979 | 0.256 |  |  |  |
| CV | 0.625 | 0.628 |  | 0.611 |  | 0.714 | 1.788 | 0.195 |  |  |  |

Figure 7.13.1 - Sole in VIIfg. Dotted lines give the length distributions of UK (England and Wales) landings; solid lines of Belgian landings












Figure 7.13.2 - Sole in VIIfg. Age composition of landings












Figure 7.13.3 - Sole in VIIfg - standardised catch proportion
Standardized catch proportion at age


Figure 7.13.4 - Sole VIIfg - Length distributions of discarded and retained fish

UK 2007
194 hauls 26 trips


UK 2008
364 hauls 62 trips


UK 2009
188 hauls 25 trips


Figure 7.13.5 - Sole VIIfg - Mean-standardised index of UK(E\&W) VIIfg Corystes survey



Figure 7.13.6 - Sole in VIIfg - Consistency plot UK-(BTS-3Q) survey


Figure 7.13.7 - Sole in VIIfg. Effort (in thousand hours, GRT corrected in case of E\&W beam trawl fleet) and LPUE (in $\mathrm{kg} / \mathrm{hour}$; or in $\mathrm{kg} / 100 \mathrm{~km}$ in case of UK(BTS-3Q) survey) for three beam trawl fleets and one survey.



Figure 7.13.8 - Sole in VIIfg - Consistency plot Uk beam trawl

log index

Figure 7.13.9 - Sole in VIIfg - Consistency plot Belgian beam trawl

## BEL-CBT


log index

Figure 7.13.10 - Sole in VIIfg. Catchability residuals for final XSA run
Residuals
Celtic Sea Sol (VIIfg) - 2010 update as


Residuals
Celtic Sea Sol (VIIfg) - 2010 update as


Figure 7.13.11 - Sole in VIIfg. Estimates of survivors from different fleets and shrinkage, as well as their different weighting in the final XSA-run

Celtic Sea Sol (VIIfg) - 2010 update assessment


Survivors


Figure 7.13.12 - Sole VIIf,g retrospective XSA analysys (shinkage SE=1.5)
Restrospective analysis
Celtic Sea Sol (VIIfg) - 2010 update as


Figure 7.13.13 Sole in VIlfg. Summary plots






Figure 7.13.14 - Sole in VIIfg. Quality control plots - Historical performance


Figure 7.13.15 - Sole in VIIfg Yield per recruit and short term forecast plots


| MFYPR version 2a |  |  |
| :--- | :--- | :--- |
| Run: Sole Vlifg_FinalYield |  |  |
| Time and date: | 14:35 |  |
| $14 / 05 / 2010$ |  |  |
| Reference point | F multiplier |  |
| Absolute F |  |  |
| Fbar(4-8) | 1.0000 | 0.1941 |
| FMax | 1.5349 | 0.2980 |
| F0.1 | 0.6615 | 0.1884 |
| F35\%SPR | 0.6693 | 0.1299 |

MFDP version 1a
Run: Sole VIlfo Fin
Run: Sole VIIfg_Fin
CELTIC SEA SOLE 2010WG COMBSEX PLUSGROUP
CELTIC SEA SOLE 2010WG COMBSEX PLUSGROUP
Time and date: 18:22 13/05/2010
Fbar age range: 4-8
Input units are thousands and kg - output in tonnes

Figure 7.13.16 Sole in VIIfg
MSY fitted stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; from the 1000 re-samples. red line median and percentiles of curves with converged estimates of Fmsy. Right hand panels :
curves plotted from the first 100 MCMC re-samples with converged Fmsy estimates.
The legends for each recruitment model show the number of converged values of FMSY


Figure 7.13.17 Sole in VIIfg
Estimates of $F$ reference points and equilibrium yield and SSB against fishing mortality using Ricker stock and recruitment model. Left hand panels : blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of Fmsy. Right hand panels : the first 100 MCMC re-samples converged Fmsy estimates. Circles show assessment estimates with the most recent year labelled.

SOLVIIFG.SUM Ricker


Figure 7.13.18 Sole in VIIfg
Estimates of $F$ reference points and equilibrium yield and SSB against fishing mortality using Smooth hockeystick stock and recruitment model. Left hand panels : blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of Fmsy. Right hand panels : the first 100 MCMC re-samples converged Fmsy estimates. Circles show assessment estimates with the most recent year labelled.

SOLVIIFG.SUM Smooth hockeystick







Figure 7.13.19 Sole in VIIfg
Fitted yield per recruit $F$ reference points, yield per recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight at age, natural mortality and maturity estimates and their c.v. Left hand panels : blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels : the first 100 re-samples.

## SOLVIIFG.SUM - Per recruit statistics



Figure 7.13.20 - Sole in VIIfg. Three year average exploitation pattern, standardised to Fbar (4-8)


### 7.14 Sole in the Southwest of Ireland (ICES Divisions VIIh-k)

## Type of assessment in 2010

No assessment was performed, however catch numbers and weighs were aggregated for the Irish landings for the years 1993-2009 and these were used to perform a yield-per-recruit analysis.

### 7.14.1 General

## Stock identity

Sole in VIIj are mainly caught by Irish vessels on sandy grounds off counties Kerry and west Cork. Sole catches in VIIk are negligible. VIIh is also considered part of the stock for assessment purposes but there is no evidence to suggest that this is actually the same stock

### 7.14.2 Data

The nominal landings are given in Table 7.14.1.
Most non-Irish landings were from VIIh which is likely to be a different stock. Because age data were only available for Irish landings (which were mainly from VIIjk) therefore the remainder of Section 7.14 concerns Irish data only in VIIgjk.

## Sampling

Figure 7.14 .1 shows that sole landings in VIIjk were mostly taken by otter trawlers in VIIj. This is reflected in the sampling.

## Data quality

Figure 7.14.2 shows the length distribution of the Irish landings in VIIjk between 1993 and 2008. Sample numbers appear to be adequate. In some years distinct modes of strong year classes are discernible but cohorts cannot easily be tracked.

Annual Age-Length-Keys (ALKs) were constructed (all quarters and gear types combined) and applied to the sampled length frequency distributions. Figure 7.14.3 shows the age distribution of sole in VIIjk between 1993 and 2009. The precision of the age distributions varies somewhat between years.

### 7.14.3 Historical stock development

Because sole in VIIh were not sampled, it would not be appropriate to raise the data to all landings in VIIhjk. Instead, the official International landings figures for VIIjk were used to raise the age distributions (Table 7.14.2).

The estimated catch numbers-at-age are given in Table 7.14.3, catch weights-at-age are given in Table 7.14.4. It is possible to track some strong and weak year classes in the catch numbers-at-age matrix. This is also illustrated by Figure 7.14.4, which shows the standardised catch proportions-at-age. Figure 7.14 .5 shows the log catch numbers-at-age. The rate of decline in catch numbers through the cohorts appears to be reasonably stable. This can be further investigated by calculating the slope of the $\log$ catch numbers $(Z)$. Figure 7.14 .6 shows the catch curve, sole under the age of 4 are not fully selected and from age 10 onwards the data get quite noisy, therefore the slope of the log catch numbers was estimated over ages 4 to 9 (Figure 7.14.7). Z estimates varied mostly between 0.2 and 0.6 .

## Yield-per-recruit

The yield-per-recruit was estimated using a method by Thompson and Bell (1934). This method requires the selectivity to be estimated. This was done by estimating the slope of the log catch numbers for ages that are fully selected and using this slope ( $Z$ ) to predict the population numbers for ages that are not fully selected. The Z was estimated on pseudo-cohorts which were standardised to take account of annual variations in the catch numbers. Figure 7.14 .8 shows that sole in VIIjk appears to be fully selected by the age of 5 and that after the age of 10 the data get very sparse. Figure 7.14 .9 shows the slope of the mean standardised log catch numbers. The predicted catch numbers from this slope were used to estimate the 'observed' selectivity. This was then modelled by applying a linear model after a logit transformation. The estimated selection curve is also shown in Figure 7.14.9. A natural mortality of 0.1 was assumed (based on the value used by the WG for sole in VIIfg) and the WG maturity ogive for sole in VIIfg was used to estimate SSB. The yield was estimated for a range of F values based on the average catch weights. Figure 7.14 .10 shows the YPR curve, $F_{\max }$ is estimated to be 0.31 . $\mathrm{F}_{0.1}$ is estimated at 0.13 . Recent (2005-2009) values of Z ranged between 0.20 and 0.35 , with $\mathrm{M}=1.0$ this would result in an F of 0.10 to 0.25 . This suggests that this stock may be within safe biological limits.

### 7.14.4 References

Thompson and Bell. 1934. W.F. Thompson and F.H. Bell, Biological statistics of the Pacific halibut fishery. 2. Effect of changes in intensity upon total yield and yield per unit of gear, Rep. Int. Fish. (Pacific Halibut) Comm. 8 (1934), p. 49.

Table 7.14.1. Sole in Divisions VII h-k (Southwest Ireland). Nominal landings (t), 1973-2009, as officially reported to ICES.

| Country | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 406 | 369 | 210 | 638 | 519 | 290 | 384 | 522 | 576 | 471 |
| Denmark | - | - | - | - | - | - |  | - | - | - |
| France | 390 | 143 | 207 | 19 | 103 | 23 | 29 | 27 | 107 | 104 |
| Ireland | 108 | 116 | 97 | 152 | 126 | 73 | 109 | 162 | 195 | 172 |
| Netherlands | 4 | 15 | 2 | 33 | 140 | 60 | - | - |  |  |
| Spain | 190 | 153 | 152 | 131 | 26 | 1 | 8 | 2 |  |  |
| UK - Eng+Wales+N.I |  | . |  |  |  |  |  |  |  |  |
| UK - England \& Wale | 6 | 5 | 24 | 11 | 12 | 11 | 18 | 42 | 83 | 108 |
| UK - Scotland | - | - | - | - | - | - | - | - | - | - |
| Total | 1104 | 801 | 692 | 984 | 926 | 458 | 548 | 755 | 961 | 855 |
| Country | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| Belgium | 411 | 474 | 318 | 442 | 271 | 254 | 252 | 353 | 358 | 312 |
| Denmark | - | - | - | - | - | - | - | - | - | - |
| France | 176 | 120 | 25 | 38 | 44 | 53 | 84 | 66 | 55 | 43 |
| Ireland | 176 | 156 | 201 | 188 | 168 | 182 | 206 | 266 | 306 | 255 |
| Netherlands | 51 | 194 | 280 | 3 |  | - | - | - | - | - |
| Spain | 38 |  |  |  |  | - | - | - | - | - |
| UK - Eng+Wales+N.I | . | . | - | . | - | $\cdot$ | 177 | 144 | 234 | 215 |
| UK - England \& Wale | 129 | 151 | 200 | 261 | 193 | 166 | . | . | . | . |
| UK - Scotland | - | - | - | - | - | - | - | - | - | 2 |
| Total | 981 | 1095 | 1024 | 932 | 676 | 655 | 719 | 829 | 953 | 827 |
| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Belgium | 317 | 338 | 433 | 375 | 368 | 346 | 101 | 8 | 13 | 154 |
| Denmark | - | - | - | - | - | - | - | - | - | - |
| France | 44 | 42 | 47 | 50 | 58 | 74 | . | 79 | 103 | 108 |
| Ireland | 237 | 184 | 243 | 183 | 203 | 221 | 207 | 111 | 125 | 130 |
| Netherlands | - | - | - | 70 | - | 7 | 1 | 10 | - | - |
| Spain | - | - | - | - | - | - | - | - | - | 1 |
| UK - Eng+Wales+N.I | 209 | 172 | 192 | 148 | 113 | 111 | 97 | 95 | 111 | 124 |
| UK - England \& Wale | . | . | . | . | . | . | . | . | . | . |
| UK - Scotland | 5 | 2 | - | - | - | - | - | - | - | - |
| Total | 812 | 738 | 915 | 826 | 742 | 759 | 406 | 303 | 352 | 517 |
| Unallocated |  |  |  | -383 | -178 | -336 | -25 | 26 | -27 | -87 |
| WG estimate |  |  |  | 443 | 564 | 423 | 381 | 329 | 325 | 430 |


| Country | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 170 | 157 | 90 | 36 | 31 | 10 | 11 |
| Denmark | - | - | - |  |  |  |  |
| France | 133 | 103 | 93 | 92 | 78 | 57 |  |
| Ireland | 105 | 111 | 98 | 63 | 78 | 72 | 71 |
| Netherlands | - | - | - | 1 |  |  |  |
| Spain | - | - | 2 |  |  | 80 | 58 |
| UK - Eng+Wales+N.I | 78 | 79 | 112 | 87 | 91 | 80 |  |
| UK - England \& Wale | - | - | - |  |  |  |  |
| UK - Scotland | - | - | - |  |  |  |  |
| Total | 486 | 450 | 395 | 279 | 278 | 219 | 140 |
| Unallocated | -241 | -160 | -69 | -7 | -1 | 6 |  |
| WG estimate | 245 | 290 | 326 | 272 | 277 | 225 |  |

Table 7.14.2. Official landings of sole in VIIjk.

| Year | Bel | Fra | Ire | Esp | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | - | 1 | 237 | . | 8 | 246 |
| 1994 | - | 0 | 176 | . | 2 | 178 |
| 1995 | - | 3 | 232 | . | 6 | 241 |
| 1996 | - | 2 | 163 | . | 1 | 166 |
| 1997 | - | 2 | 187 | . | 2 | 191 |
| 1998 | - | 9 | 208 | . | 2 | 219 |
| 1999 | 96 | 0 | 199 | . | 1 | 296 |
| 2000 | 8 | 6 | 103 | . | 0 | 117 |
| 2001 | 7 | 13 | 114 | - | 0 | 134 |
| 2002 | 69 | 23 | 121 | . | 0 | 213 |
| 2003 | 48 | 20 | 82 | . | 0 | 150 |
| 2004 | 2 | 7 | 78 | . | 0 | 87 |
| 2005 | - | 7 | 70 | <0.5 | 0 | 77 |
| 2006 | - | 11 | 49 | - | 1 | 61 |
| 2007 | - | 9 | 74 | - | 0 | 83 |
| 2008 | - | 8 | 69 | - | 0 | 77 |
| 2009* | 0 | ** | 68 | - | 0 | 68 |

* Preliminary data
** Not available at the time of the Working Group.

Table 7.14.3. Catch numbers-at-age for sole in VIIjk.

|  |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 33 | 218 | 224 | 77 | 56 | 57 | 32 | 21 | 12 | 11 | 5 | 5 | 14 |
| 1994 | 23 | 117 | 130 | 69 | 41 | 22 | 19 | 11 | 12 | 13 | 11 | 4 | 27 |
| 1995 | 0 | 279 | 81 | 174 | 117 | 51 | 15 | 15 | 4 | 22 | 8 | 8 | 6 |
| 1996 | 12 | 46 | 116 | 80 | 53 | 54 | 31 | 8 | 5 | 6 | 10 | 3 | 33 |
| 1997 | 39 | 161 | 84 | 110 | 43 | 41 | 38 | 16 | 1 | 0 | 4 | 3 | 17 |
| 1998 | 23 | 137 | 113 | 59 | 93 | 40 | 43 | 34 | 9 | 5 | 3 | 5 | 32 |
| 1999 | 51 | 179 | 218 | 187 | 67 | 77 | 30 | 28 | 19 | 2 | 11 | 1 | 19 |
| 2000 | 39 | 96 | 83 | 42 | 29 | 16 | 21 | 11 | 17 | 8 | 3 | 0 | 5 |
| 2001 | 65 | 115 | 53 | 49 | 38 | 22 | 22 | 14 | 9 | 4 | 2 | 5 | 8 |
| 2002 | 13 | 139 | 183 | 66 | 38 | 39 | 15 | 8 | 24 | 8 | 21 | 5 | 31 |
| 2003 | 2 | 54 | 93 | 128 | 76 | 45 | 18 | 4 | 5 | 9 | 14 | 0 | 9 |
| 2004 | 7 | 18 | 92 | 48 | 36 | 19 | 14 | 6 | 8 | 1 | 7 | 1 | 20 |
| 2005 | 10 | 34 | 47 | 65 | 17 | 38 | 21 | 9 | 4 | 4 | 0 | 4 | 14 |
| 2006 | 13 | 29 | 30 | 28 | 38 | 18 | 16 | 11 | 6 | 4 | 1 | 1 | 11 |
| 2007 | 1 | 44 | 36 | 30 | 44 | 42 | 21 | 16 | 10 | 4 | 4 | 1 | 8 |
| 2008 | 1 | 25 | 90 | 42 | 21 | 20 | 25 | 11 | 8 | 5 | 3 | 3 | 7 |
| 2009 | 0 | 15 | 38 | 75 | 31 | 17 | 16 | 16 | 6 | 6 | 5 | 1 | 4 |

Table 7.14.4. Catch weight-at-age for sole in VIIjk.

| $\mathbf{2}$ |  | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1 9 9 3}$ | 0.154 | 0.221 | 0.275 | 0.342 | 0.412 | 0.455 | 0.511 | 0.496 | 0.628 | 0.567 | 0.761 | 0.499 | 0.706 |
| 1994 | 0.143 | 0.233 | 0.278 | 0.346 | 0.421 | 0.453 | 0.514 | 0.552 | 0.610 | 0.632 | 0.632 | 0.583 | 0.737 |
| 1995 |  | 0.194 | 0.322 | 0.362 | 0.338 | 0.370 | 0.493 | 0.452 | 0.722 | 0.579 | 0.401 | 0.297 | 0.592 |
| 1996 | 0.138 | 0.169 | 0.230 | 0.307 | 0.435 | 0.421 | 0.505 | 0.587 | 0.613 | 0.711 | 0.755 | 0.643 | 0.698 |
| 1997 | 0.133 | 0.200 | 0.281 | 0.334 | 0.409 | 0.526 | 0.618 | 0.592 | 0.679 |  | 0.692 | 0.846 | 0.922 |
| 1998 | 0.137 | 0.223 | 0.281 | 0.357 | 0.379 | 0.448 | 0.515 | 0.554 | 0.455 | 0.646 | 0.497 | 0.641 | 0.805 |
| 1999 | 0.152 | 0.192 | 0.308 | 0.345 | 0.400 | 0.426 | 0.461 | 0.575 | 0.578 | 0.657 | 0.449 | 0.896 |  |
| 2000 | 0.180 | 0.210 | 0.255 | 0.396 | 0.416 | 0.472 | 0.502 | 0.489 | 0.505 | 0.452 | 0.554 |  | 0.641 |
| 2001 | 0.164 | 0.228 | 0.295 | 0.337 | 0.394 | 0.481 | 0.548 | 0.530 | 0.587 | 0.795 | 0.542 | 0.740 | 0.727 |
| 2002 | 0.203 | 0.198 | 0.255 | 0.305 | 0.470 | 0.490 | 0.473 | 0.655 | 0.732 | 0.724 | 0.627 | 0.616 | 0.895 |
| 2003 | 0.168 | 0.191 | 0.296 | 0.323 | 0.329 | 0.378 | 0.371 | 0.575 | 0.503 | 0.548 | 0.477 |  | 0.600 |
| 2004 | 0.095 | 0.200 | 0.198 | 0.294 | 0.313 | 0.353 | 0.287 | 0.581 | 0.632 | 0.498 | 0.595 | 0.498 | 0.724 |
| 2005 | 0.128 | 0.168 | 0.198 | 0.249 | 0.383 | 0.318 | 0.340 | 0.445 | 0.525 | 0.468 |  | 0.489 | 0.614 |
| 2006 | 0.160 | 0.180 | 0.205 | 0.257 | 0.298 | 0.354 | 0.354 | 0.377 | 0.456 | 0.377 | 0.612 | 0.438 | 0.718 |
| 2007 | 0.154 | 0.208 | 0.268 | 0.282 | 0.329 | 0.341 | 0.378 | 0.395 | 0.449 | 0.376 | 0.418 | 0.554 | 0.522 |
| 2008 | 0.143 | 0.205 | 0.236 | 0.275 | 0.305 | 0.339 | 0.339 | 0.395 | 0.389 | 0.448 | 0.559 | 0.450 | 0.631 |
| 2009 | 0.123 | 0.196 | 0.234 | 0.265 | 0.268 | 0.318 | 0.386 | 0.420 | 0.393 | 0.417 | 0.368 | 0.476 | 0.587 |

sol-7h-k
Relative landings and number of samples


Figure 7.14.1. Irish Operational landings and sampling levels (number of samples) for sole in VIIjk by quarter (top), geartype (middle) and ICES Division (bottom). The sampling appears to be representative of the landings.


Figure 7.14.2. Length frequency distribution of the Irish landings of sole in VIIjk between 1993 and 2009. All gears and quarters combined. Sampling was poor during 2006 and 2007.


Figure 7.14.3. Age distribution of sole in VIIjk between 1993 and 2009. All gears and quarters combined.

Sole VIInk
Standardised catch proportions-at-ag


Figure 7.14.4. Standardised catch proportions-at-age for sole in VIIjk. Grey bubbles represent higher than average catch-at-age and black bubbles represent lower than average catch-at-age.

## Sole VIIhk

Log catch numbers


Figure 7.14.5. Log catch numbers-at-age (ages 4-8).


Figure 7.14.6. Catch curve of plaice in VIIbc. Plaice from the age of 4 appear to be fully selected; the data get quite noisy from the age of 7 onwards.


Figure 7.14.7. Z estimated over pseudo-cohorts as the slope of the $\log$ catch numbers.


Figure 7.14.8. Selectivity was modelled by fitting a line through the mean $\log$ standardised catch numbers of ages 4 to 14 to predict the expected catch numbers for ages 1 to 3 if these were fully selected. The proportions of observed divided by expected catch number were taken as the 'observed' selectivity. This was then modelled using a logit transformation.


Figure 7.14.9. YPR analysis using the Thompson-Bell approach. Recent estimates of $Z$ were between 0.2 to 0.5 which translates to an F of 0.1 to 0.4 .

### 7.15 Whiting in Division VIIe-k

Type of assessment in 2010
Update assessment. Same Advice as Last Year.
ICES advice applicable to 2009 and 2010
Exploitation boundaries in relation to precautionary limits: The current estimates of fishing mortality and SSB are uncertain, but SSB shows a decreasing trend while recruitment has been low in recent years although the 2007 year class is above average, and the 2008 year class may be very strong. In order to reverse the trend in SSB, ICES considers that fishing mortality should be reduced. However, ICES cannot quantify the required reduction in fishing mortality.

In addition, ICES offers the following consideration: surveys indicate that the 2007 year class is above average, and the 2008 year class may be very strong. Management measures should be introduced in the Celtic Sea to reduce discarding of these year classes in order to maximize their contribution to future yield and SSB.

### 7.15.1 General

## Stock description and management units

The TAC for whiting is set for Divisions VIIb-h and VIIk. However VIIj has been omitted from the area for the last three years. This assessment area does not correspond to the TAC area. Whiting in VIIb,c are not assessed and whiting in VIId are included in the WDNSSK assessment of the North Sea stock. Any management measures implemented for this stock should be consistent with the assessment area.


Red Boxes-TAC/Management Areas Blue Shading- Assessment Area
The 2010 TAC for whiting VIIb-h and k has been reduced from 16949 t to 14407 t . This TAC has not been considered restrictive, with officially reported VIIe-k landings totalling 3270 t in 2009, although this does not include French landings, around 3000 t in 2008. The assessment is based on landings only, as reported in logbooks, and does not include discards. The introduction of buyers and sellers legislation in 2007 should improve landings statistics.

TAC in 2009

| Species: Whiting <br> Merlanglus merlangus |  | Zone | VIIb, VIIc, VId, VIIe, VIIf, VIIg. VIIh and VIk (WHG/7X7A.) |
| :---: | :---: | :---: | :---: |
| Belgium | 163 |  |  |
| France | 9999 |  |  |
| Ireland | 4918 |  |  |
| The Netherlands | 81 |  |  |
| United Kingdom | 1788 |  |  |
| EC | 16949 |  |  |
| TAC | 16949 |  | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 applies. <br> Article 4 of Regulation (EC) No 847/96 applies. <br> Article 5(2) of Regulation (EC) No 847/96 applies. |

TAC in 2010

| Species:Whiting <br> Merlangius merlangus | Zone:VIIb, VIIc, VIId, VIle, VIlf, VIlg, VIIh and VIlk <br> (WHG/7X7A.) |  |
| :--- | :--- | :--- | :--- |
| Belgium | 133 |  |
| France | 8180 |  |
| Ireland | 4565 |  |
| The Netherlands | 66 | Analytical TAC |
| United Kingdom | 14463 |  |
| EU | 14407 |  |
| TAC |  |  |

## Fishery in 2009

ICES officially reported landings for Divisions VIIe-k and landings as used by the Working Group are given in Table 7.15.1. It was not possible to compare the ICES officially reported landings and those reported to the WG for 2009 due to the lack of official French landings. In 2008 higher WG landings were reported by France. Official landings were wrongly allocated to VIIbc, which has now been corrected reducing the discrepancy for 2008. In 2009 international landings provided to the Working Group are very similar to those of 2008.

Minimal revisions ( $<1 \mathrm{t}$ ) to 2008 landings were submitted to the WG. ICES Official landings increased by $\sim 1200 \mathrm{t}$, primarily resulting from French revisions. Landings from Spain, UK Scotland and the Channel Islands have also now been reported (combined <50 t).

The VIIe-k whiting stock is primarily targeted by otter trawlers and to a lesser extent Scottish seines and beam trawls. Otter trawlers utilize two mesh size ranges to $70-89 \mathrm{~mm}$ and 100-119 mm. Effort of trawlers utilizing these two mesh size ranges has remained relatively stable within the Celtic Sea as a whole, however effort of the larger mesh range has declined within VIIf and VIIg over recent years. The vessels utilising these mesh ranges have different species selectivity patterns. Several main species groups are tar-
geted by otter trawlers catching whiting, as part of a targeted mixed gadoid fishery and as bycatch within the Nephrops and hake, anglerfish, and megrim fisheries. Beam trawlers operate to the eastern side of the assessment area, VIIe-h where small quantities of whiting are taken as a bycatch species in flatfish, anglerfish, and ray target fisheries. The spatial distributions of landings by country in 2009 are given in Figure 7.15.1. Irish catches are primarily from within VIIg particularly within 31E2 and 31E3. Landings also originate, to a lesser extent from VIIj. In previous years French landings have exhibited similar spatial and temporal focus around 31E3. No French spatial data was available for 2009. The majority of UK landings are from otter trawlers in VIIe, and focused within 29E5 and 29 E 6.

### 7.15.2 Data

An overview of the data provided and used by the WG is provided in Table 2.1.

## Landings

National landings and numbers-at-age data were aggregated for the area VIIe-k following methodology described in the Stock Annex, with the exception of the French data. Due to problems with the French logbook database, the landings data were considered incomplete and not available by quarter and métier. Although considered incomplete the available data, around 2100 t , was included in the assessment. The landings data was allocated to quarters using the mean proportion by quarter over the period 2006-2008, which appeared to be reasonably stable. Secondly, the sample length distributions within each quarter were assumed to be representative of the landings of each métier. National sampling levels for the landings are presented in Table 2.1.

The length compositions from various fleets for 2009 are displayed in Table 7.15.2 and Figure 7.15.2. The landings length distributions of the Irish otter trawl, UK and French fleets, which account for the majority of the landings, are similar, peaking around 3234 cm . Scottish seine fleets land a wider distribution reaching sizes over 50 cm . The peak length ranges from 37 cm to 44 cm , with a slight tendency for seiners in VIIg to land smaller fish than in VIIj.

The international catch numbers-at-age are given in Table 7.15.3 and Figure 7.15.3. It is possible to track strong year classes in the landings-at-age matrices. The age distribution has remained similar over time, with the exception of periods where strong year classes pass through older ages. Age group 0 was included in the assessment data to allow inclusion of 0 -group indices in the XSA, although landings at this age were not recorded in most years. Very small landings of 0 -group whiting were not included in the catch-at-age data-file to avoid spurious F-shrinkage effects at this age. Mean weights-at-age in the catch and stock (Tables 7.15.4 and 7.15.5) were derived as per the methodology described in the Stock Annex. The stock weights are shown in Figure 7.15.4. There is some variability in stock weights particularly at older ages. There is some indication of a decreasing trend in weights for ages 6 and 7 over the whole time period.

## Discards

Discard data are available from the Irish fishery since 1994 (ICES: SGDBI, 2002), from French sampling in 1991, 1997, and 2005-2009, and for the UK (E\&W) fisheries from 2001-2009. These data are not used in the assessment as the data available does not cover the full time-series of landings-at-age-data, and historically sampled fleets may not be
representative of the main fleets involved in the fishery. Furthermore, there is a need to examine and agree the best raising practice for the various fleets. Discard rates are substantial ( $>50 \%$ by fleet/quarter) and variable. It is not clear if current sampling intensity will obtain precise enough annual estimates to support an assessment method where catch numbers are assumed to be exact as in XSA.

A summary of the 2009 discard sampling and discard rates is presented in Table 7.15.6. Discard rates between years, quarters and fleets can be very variable, although the UK data are similar to the range observed in 2008. Discarding is much higher for Irish otter trawls in VIIg than last year.

Discarded whiting length distributions from 2009 Irish and French otter trawlers, and all UK gears were made available to the WG (Figure 7.15.5). The available data indicate that discarding occurs above the 27 cm MLS with some fish being discarded up to 50 cm in some fleets. The discard Lso's for most countries/fleets is around $25-27 \mathrm{~cm}$.

Age compositions for Irish discard data were provided for otter trawlers in VIIg and VIIj for 2003-2009 indicating discarding from age 0 up to age 8 in some years. Substantial discarding of ages 1 and 2 occurs for most years (Figure 7.15.6). Discard numbers-at-age have not yet been calculated for other fleets.

## Biological

Mean stock weights- and numbers-at-age data were calculated, following methodology described in the Stock Annex.

Natural mortality was assumed to be 0.2 over all age groups and years.
Available data on maturity-at-age are described in the Stock Annex. Since 2006 the knifeedge maturity ogive has been replaced with indices calculated based on data from the UK WCGFS but a fixed vector is still used. Recent maturity sampling by Ireland and the UK on dedicated surveys confirms the use of this ogive but is in-sufficient to provide annual data.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5 +}$ |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Maturity | 0 | 0.39 | 0.90 | 0.99 | 0.99 | 1.00 |

The proportions of F and M before spawning were both set to zero to reflect the SSB calculation date of January 1st.

## Surveys

A time-series of available standardized survey abundance indices for ages $0-3$ are displayed in Table 7.15.8. Further details of these surveys are given in WGSSDS 2008 Table 1.3.3 and described in the Stock Annex. Figure 7.15 .9 shows standardised and $\log$ standardised abundance indices by age ( $0-3$ ) for the three surveys used in the assessment by year class. In total four fishery-independent survey indices including 2009 data were available to the WG. The strong 1999 year class is evident in all surveys. The complete time-series and ages available from these surveys are given in the tuning fleet information available to the Working Group (Table 7.15.8).

The internal consistency of the surveys was examined, pairwise scatterplots of log num-bers-at-age were examined, bearing in mind that the correlations may be impacted by changed in fishing mortality. Plots for the three surveys included in the assessment are provided in Figure 7.15.7b. Year effects were examined with mean log standardized plots of indices by age and year (Figure 7.15.8a). Cohort tracking was examined with mean log standardized plots of indices by age and cohort (Figure 7.15.8b).

The FR-EVHOE survey log indices scatterplots display reasonable positive correlation between adjacent ages. The mean log standardized indices by year display a year effect in 2006 and by cohort demonstrates good tracking of stronger year classes. The UK-WCGFS also demonstrated inconsistencies between years in the log-index scatterplots but reasonably consistent catch-curves. Log-indices for the Irish VIIg swept-area survey reveal some positive correlation for younger ages. The mean log standardized index by year demonstrated some slight year effect in 2003 which was the first year of the new series.

## Commercial Ipue

Estimates of commercial lpue, from 1995 to 2009, were available for the Irish otter trawl, Scottish seine, and beam trawl fleets operating in Divisions VIIg and VIIj (Table 7.15.9 and Figure 7.15.10). The effort-series is raw effort in hours uncorrected for changes in vessel power or changes in species targeting (i.e. métier compositions). Increased Irish VIIg otter trawl landings and lpue occurred 2005-2007, then returned to prior levels. This increase coincides with the 1999 year class passing through the fishery. Landings and effort for this fleet increased in 2009, although little change in lpue is observed. The recent elevated effort has been associated with the displacement, and subsequent relocation of effort in response to restrictive management in other areas particularly VIa and VIIa. The VIIj otter trawl fleet landings, effort, and lpue show similar levels since 2005, although slight increases to those of 2008 are observed. In the earlier part of the time-series lpue for the IR-7G-SSC and IR-7J-SSC showed declining tends. Since 2006/2007 lpue has increased. Landings by these two fleets however are low. Effort and lpue data for the Irish beam trawls (TBB) operating in VIIg and VIIj are also included in Table 7.15 .9 but is not plotted as landings, effort and lpue are minimal.

Estimates of commercial lpue, up to 2008 were available for French gadoid trawlers and French Nephrops trawlers operating in Divisions VIIf,g (Table 7.15.9 and Figure 7.15.10). Fishing effort in the FR-GADOID fleet has been declining since 1989, while the effort in the FR-NEPHROPS has declined since 1992. The FR-GADOID fleet's lpue increased to high levels in 1994 and 1995 but declined since. Sharp increases in lpue for the French gadoid fleet occurred in both 1998 and 2005, since which lpue has declined. Lpue for the FR-NEPHROPS fleet peaked in the mid-to-late 1990s, having declined since to levels similar to the early 1980s. Landings, effort and lpue for both these fleets currently demonstrate the lowest levels within the time-series. Limited lpue data from France are available for Divisions VIIj-k, but they are not considered representative. The commercial tuning fleets available to the assessment are given in Table 7.15.8.

Abundance indices-at-age were available for three commercial fleets, the French gadoid, and Nephrops fleets, and the Irish otter trawl fleet. As with the surveys, the internal consistency of these fleets (Figure 7.15.7a), any year effects (Figure 7.15.8a) and cohort tracking (Figure 7.15.8b) were examined. The French commercial Nephrops index demonstrates very good internal consistency. The French gadoid fleet shows good consistency, although consistency at age 3 is slightly poorer. The IROTB-7g\&j previously used in the
assessment was not considered as a consequence of poor cohort tracking and a priori concerns about changes in targeting practice and fishing power because of recent fleet changes since 2002.

## Other relevant data

Meetings held with representatives of the fishing industry raised no specific concerns or comments.

### 7.15.3 Historical stock development

An XSA assessment was carried out for this stock applying the same settings as last year's update assessment, with the addition of 2009 data. The settings previously used and applied this year are detailed within the Stock Annex.

## Data screening

The general methodology is outlined in Section 2. Preliminary investigations were carried out using FLR under R version 2.4.1. The packages FLCore 1.4-3, FLAssess 1.4.1, FLXSA 1.4-2 and FLEDA 1.4-2 were used.

## Final update assessment

The assessment was carried out with FLXSA 1.4-2 under R version 2.4.1. The assessment uses the same settings as last year (detailed below), with the exception of the French commercial tuning fleets which were not updated in 2009 due to data non-availability. The tuning data available, and the subset used in the assessment, are given in Table 7.15.8. No exploratory runs were carried out for this assessment.

|  |  | 2009 | 2010 |
| :---: | :---: | :---: | :---: |
| Catch date range: | Years | 82-08 | 82-09 |
|  | Ages | 0-7+ | 0-7+ |
| Fbar Age Range: |  | 2-5 | 2-5 |
| Assessment Method: |  | XSA | XSA |
| Commercial Tuning Fleets: |  |  |  |
| FR-Gadoid Late | Yrs | 93-08 | 93-08 |
|  | Ages | 3-6 | 3-6 |
| FR-Nephrops | Yrs | 93-08 | 93-08 |
|  | Ages | 3-6 | 3-6 |
| Survey Tuning-series: |  |  |  |
| FR-EVHOE | Yrs | 97-08 | 97-09 |
|  | Ages | 0-4 | 0-4 |
| UK-WCGFS | Yrs | 87-01 | 87-01 |
|  | Ages | 1-6 | 1-6 |
| IR-IGFS Swept area | Yrs | 99-08 | 99-09 |
|  | Ages | 0-6 | 0-6 |
| Time taper: |  | No | No |


|  |  | 2009 | 2010 |
| :--- | :--- | :---: | :---: |
| Q plateau age: |  | 5 | 5 |
| F shrinkage S.E: | Num yrs | 1 | 1 |
|  | Num ages | 5 | 5 |
|  |  | 3 | 3 |
| Fleet S.E: | 0.5 | 0.5 |  |

The full XSA diagnostics are given in Table 7.15.10. Substantially higher survivor estimates are given by the IR-IGFS Swept area survey than the FR-EVHOE survey for the 2009 year class (age 0), although weighting between the two is almost equal. The FREVHOE survey estimated substantially higher survivors for both the 2008 year class (age 1) and the 2007 year class (age 2). The two estimates of survivors converge from the 2005 year class (age 4). The French gadoid fleet generally gave higher estimates than the Nephrops tuning-series. Figure 7.15 .11 shows the scaled weights received by each fleet in the assessment.

The log-catchability residuals from the XSA fit are plotted for each tuning-series in Figure 7.15.12. There are some year effects and noise in the short time-series of data. Year effects can be seen within the French commercial fleets in recent years. The Gadoid fleet displays increased catchability while the Nephrops fleet reveals reduced catchability suggesting a shift in the fishing patterns of this two fleets and noise in the short time-series of data.

The retrospective pattern is shown in Figure 7.15.13. The retrospective bias around the 1999 year-class remains, since which it has been relatively consistent. Recruitment bias is a result of the non-inclusion of discards in the assessment while discarding rates are high. The large 2008 recruitment given by the assessment has been revised downwards. The bias in F of recent years is not seen in the last two assessments. SSB shows little bias in recent years.

Estimates of fishing mortality and stock numbers from the final XSA are given in Tables 7.15.11 and 7.15.12. These are summarized in Table 7.15.13 and Figure 7.15.14. The assessment this year reveals a continued decline in fishing mortality. Although the last two years of recruitment have been revised downward they remain above average of recent years. Recruitment of 2009 is below the time-series average.

## Comparison with previous assessments

This assessment is an update of the assessment settings carried out since 2007, with the exception of the French commercial tuning fleet for which 2009 data was not available. Minor revisions to landings and landings numbers-at-age have been included. The current assessment estimates of F agree with those estimated last year, as do SSB with a slight reduction in 2008. This is coupled with a decreased estimate of recruitment. This implies revisions are due to the assessments ability to predict recruitment, opposed to changes in landings inputs.

## State of the stock

Trends in landings, $\mathrm{F}(2-5)$, SSB, and recruitment are presented in Table 7.15.13 and Figure 7.15.14. SSB displays peak biomass in the mid-1990s following a series of good recruitment in preceding years. Subsequently SSB has shown a declining trend, which was
temporarily halted by the strong 1999 year class. SSB for the last two years shows an increase, particularly in 2009, estimated to be 33680 t , well above $\mathrm{B}_{\mathrm{pa}}(21000 \mathrm{t}$ ). Fishing mortality ( $\mathrm{F}_{\mathrm{bar}}$ ) is estimated to have declined in the last two years following a period of increase which peaked in 2007. The 2009 estimate of fishing mortality is estimated to be 0.48 .

Recruitment estimates of 2007 and 2008 year classes are above those of the six preceding year classes. The very large estimate of the 2008 year class has been revised downward by the current assessment. The WG believe this to be an above average year class, although the size is still considered to be uncertain. The relationship between SSB and recruitment is poorly defined (Figure 7.15.16) and there is no evidence of reduced recruitment at lower levels of SSB.

### 7.15.4 Short-term projections

## Estimating year-class abundance

The XSA estimate of the 2007 and 2009 year classes ( 71.9 m and 56 m respectively) were kept for the prediction. The 2007 estimate has demonstrated only a minor downward revision from last year. The 2009 year class has been retained within the forecast as it is similar to the recruitment of the early 2000s. The 2008 year-class XSA estimate has been reduced by $25 \%$ at age 0 and age 1 adjusted accordingly. This should account for the likely hood of further downward revisions of initial year-class size as seen in the retrospective analysis of the 1999 year class. Subsequent year classes have been set at GM recruitment over the full time-series (using the adjusted 2008 recruitment) of 69136.

The Working Group estimates of year-class strength can be summarized as follows (re-cruitment-at-age 0):

| Year class | Thousands | Basis | FR-EVHOE | IR-GFS7gSwept | Shrinkage |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 71912 | XSA | $36.3 \%$ | $56.9 \%$ | $6.8 \%$ |
| 2008 | 89231 | XSA-25\% | $35.8 \%$ | $55.5 \%$ | $8.8 \%$ |
| 2009 | 56010 | XSA | $45.1 \%$ | $54.9 \%$ |  |
| Onwards | 69136 | GM 1982- |  |  |  |
|  |  | 2009 |  |  |  |

## Short-term predictions

Input data for the predictions are given in Table 7.15.14. The exploitation pattern is based on the fishing mortalities averaged over 2007-2009, unscaled ( $\mathrm{F}_{2-5}=0.782$ ). Weights-at-age were the mean of 2007-2009. Table 7.15 .15 is the management option table and Table 7.15.16 gives the detailed results. Figure 7.15.20 gives the short-term yield and SSB forecasts.

Assuming status quo F, landings are predicted to be around 14472 t in 2010 and 15281 t in 2011 (Table 7.15.16). SSB is predicted to be at 42731 t in 2010, above $\mathrm{B}_{\mathrm{pa}}$, declining in 2011 to 41408 t, and in 2012 to 39415 t . This is in contrast to last year's forecast which indicated SSB would increase. Estimates of the relative contribution of recent year classes to the 2011 landings and 2012 SSB are displayed in Table 7.15.17. The assumed GM recruitment accounts for $1.4 \%$ of the landings in 2011 and $42.3 \%$ of the SSB in 2012.

### 7.15.5 Medium-term projection

No medium-term projections were carried out.

### 7.15.6 Maximum Sustainable Yield Evaluation

The whiting VIIe- k time-series of assessment stock and recruit estimates, fishing mortal-ity-at-age (average of the most recent three years, with CVs from eight year average of standardised F to allow greater variation), catch and stock weights (eight year averages), maturity and natural mortality-at-age were used to estimate proxies for the fishing mortality biomass and landings at maximum sustainable yield ( $\mathrm{F}_{\mathrm{msy}}, \mathrm{B}_{\mathrm{msy}}$ and MSY) within the srmsymc ADMB programme. The input data files (sen and sum) are presented in Tables 7.15.18 and 7.15 .19.

Three stock and recruit models are fitted by the programme, Ricker, Beverton and Holt and the smooth hockey stock, a summary of which is given in Figure 7.15.17. No iterations could be fitted to the Beverton-Holt recruitment model, and therefore discounted from further consideration for this stock. Figures 7.15 .18 a and 7.15 .18 b show output plots for the remaining two models. Based on the A.I.C. all models have an equal fit to the available data (Table 7.15.20). Table 7.15 .20 also details the estimates of $\mathrm{F}_{\text {crash, }} \mathrm{F}_{\text {msy }}, \mathrm{B}_{\text {msy }}$ and MSY, presented with their percentiles and coefficients of variation.

The stock-recruitment relationship for this stock is poorly defined with little information in the data to inform the shape of the stock-recruitment relationship. There is a high rejection rate of iterations for both the Ricker and hockey stick models. The F parameters are poorly defined by the models having both wide distributions and high levels of uncertainty. Furthermore, the form of the YPR curve is poorly determined. Figure 7.15.19 illustrates the uncertainty in yield-per-recruit curve. The estimates from the YPR are presented in Table 7.15.20 where high CV's are observed particularly for $\mathrm{F}_{\text {max. }}$.

There are two important considerations to be noted in addition to the evaluation of the MSY analysis.

- This assessment is accepted as trends only due to high levels of uncertainty within the underlying data, the outputs of which are required to determine MSY parameters.
- A large amount of potential yield is lost through discarding. The yield in this analysis refers to landings only, whilst discarding for this stock is high.

The Working Group is unable to provide values of $\mathrm{F}_{\text {msy }}$ as a result of the above considerations. However, the Working Group considers it likely that fishing mortality is above Fmsy and a reduction of discarding is needed to improve the catch selectivity pattern.

## Yield-per-recruit analysis

Results of a yield-per-recruit analysis, using MFYPR, (Table 7.15.21 and Figure 7.15.20) indicate that $\mathrm{F}_{\max }$ is $0.73,86 \%$ of status quo F (0.78). Assuming $\mathrm{F}_{0.1}$ (F 0.17 ), the current exploitation pattern, and the GM recruitment applied within the short-term forecast, longterm yield is estimated to be 11490 t below the current TAC and an SSB of 82846 t . Maintaining status quo F would yield 13668 t with an SSB of 39221 t .

### 7.15.7 Biological reference points

Precautionary approach to reference points.
The Working Groups current approach to reference points is outlined in Section 2. A summary of reference point proposals to date and their technical basis is given in the Stock Annex. The reference points were not re-examined in this update assessment, those currently adopted and their basis are as follows:

| FLIM | No Proposal |
| :--- | :--- |
| FPA | No Proposal |
| $B_{\text {LIM }}$ | 15000 t |
| B $_{\text {PA }}$ | 21000 t |

### 7.15.8 Management plans

No management plan has been agreed or proposed.

### 7.15.9 Uncertainties and bias in assessment and forecast

## Sampling

The sampling levels for those countries supplying data for 2009 are given in Table 2.1. Sampling levels of the landed catch for recent years are considered to be sufficient to support current assessment approaches. Sampling levels were not available by fishery/métier and the WG was therefore unable to evaluate whether or not current sampling levels are sufficient to support fishery/métier disaggregated assessment approaches.

## Ageing

The strong recent cohorts passing through the fishery indicates that age estimation is consistent throughout the age range used in the assessment, although some underestimation does occur at older ages.

## Discards

Discarding is a major feature of most fisheries catching whiting in the Celtic Sea. The non-inclusion of discard data in the assessment could explain some of the retrospective bias problems and changing catchabilities in commercial fleets observed throughout the assessment period. The availability of discard data has improved in the most recent years since the implementation of the DCF sampling programmes.

## Surveys

Currently, there are two IBTS surveys (French and Irish) covering the Celtic Sea. Although these surveys normally catch large quantities of whiting they seem prone to year effects as has been observed for this species in other areas (e.g. Irish Sea, North Sea). These surveys give very different estimates of the 2009, 2008, 2007, and 2006 year classes. The estimation of younger year classes is one of the most important factors in the shortterm development of the stock.

## Misreporting

The level of misreporting of this stock is not known and underreporting has previously been considered unlikely to have been a significant source of unaccounted mortality of whiting in the assessment because the TAC has been in excess of recent landings.

## Consistency

Inter-annual comparison between the results of this year's and last year's assessments shows consistent estimates up until 2005. Estimated recruitment for the 2007 and 2008 year class have been revised downward, by $18 \%$ and $55 \%$ respectively. Estimates of F are highly consistent between assessments, only a slight downward revision is observed in 2007 (<2\%). SSB estimates exhibit a small downward revision (11\%).

SSB has been rescaled upwards slightly in the past when the full time-series of commercial tuning data was included in the assessment. Consistency between more recent assessments showed some problems with recruitment and SSB estimates as strong year classes during the 1990s passed through the fishery and were heavily discarded. Assessments for the last few years have been reasonably consistent for SSB with some downward revisions, while F is revised upwards. Estimates of recruitment in the most recent assessment, remains problematic.

### 7.15.10 Recommendation for next Benchmark

The 2009 assessment was accepted for trends only by the Celtic Sea Review Group which had no specific comments on the assessment of whiting VIIe-k. The RG comment that the WG should provide potential management actions that can be taken to protect the 2008 year class.

A benchmark assessment of whiting is necessary.
Problem: The assessment of this stock has not been accepted for a number of years and considered to be indicative of trends only. The primary uncertainty of this assessment is underestimation of mortality. Currently the assessment is based on landings only. Discarding is a major feature of most fisheries catching whiting in this stock area. Mortality may therefore be grossly underestimated in younger ages. This could explain some of the retrospective bias problems and changing catchabilities in commercial fleets observed throughout the assessment period.

Solution: The available discard data has improved in the most recent years since the implementation of the DCR sampling programmes. Data are now available for the main fleets, operating within VIIe-k. Work is now required to compile a complete time-series of discard data, and evaluate raising options and uncertainty levels. Assessment model and settings then need to be reviewed to ensure optimum performance.

Year of last benchmark: No benchmark assessment of this species has been carried out. Exploratory analyses were carried out in the WGSSDS up until 2007.

WGCSE 2011 should review the time-series of discard data and options for inclusion of into this assessment. Until this happens WGCSE will not propose a time frame for the next Benchmark.

Expertise required: Expertise in discard raising and uncertainty methods, in addition to expertise in assessment methods permitting inclusion of discard data.

A further matter for consideration is the improvement of commercial tuning fleets by selection of vessel subsets with consistent spatial and temporal effort and catch composition over the majority of the time-series, moving towards the métier based approach. This would require a detailed analysis of vessel behaviour.

### 7.15.11 Management considerations

Catches and SSB in VIIe-k whiting fluctuate considerably depending on year-class strength. Indications are that the 2008 year class is strong and the 2007 and 2009 are stronger than recent years. Management measures should be considered to reduce discarding of this year class such that yield and SSB contributions can be maximized. This could be achieved through gear modifications to increase the likelihood of small whiting passing through the gear, such as introduction of larger minimum mesh sizes, separator panels, or grids.

Technical measures applied to this stock include a minimum landing size 27 cm ) and minimum mesh sizes applicable to the mixed demersal fisheries. These measures are set depending on areas and years by several regulations. Whiting are caught in directed gadoid trips and as part of mixed fisheries throughout the Celtic Sea, as well as bycatch within Nephrops fisheries. Discard rates are high as a consequence of the low market value of the species, particularly at smaller sizes. High-grading above the MLS to some extent is also prevalent in most fisheries. The current assessment doesn't include discard estimates. Recent selection data from FTFB should be investigated at the next Benchmark Workshop.

From the 1 February to the 31 March fishing activity has been prohibited within ICES rectangles: 30E4, 31E4, 32E3 (excluding within six nautical miles from the baseline) annually since 2005 to protect the cod stock. The impact of this on whiting remains unclear but spatial distribution of landings in 2009 suggest that landings from the closed rectangles are lower than those of adjacent rectangles. Irish quarterly landings by rectangle indicate little or no landings from within these closed rectangles during the first quarter.

There have been major changes in fleet dynamics over the period of the assessment. Effort in the French gadoid fleet has been declining since 1999. Irish otter trawl effort in VIIg,j has been stable over the last four years. During this period there has been a fleet modernisation and several decommissioning schemes in Ireland both within the national whitefish fleet and beam trawl fleet. The most recent round of decommissioning occurred in 2008 and 2009 removed 40 vessels which had operated within the Celtic Sea in 20072008. The decommissioned vessels accounted for $15-16 \%$ of whiting landings from the stock area in 2007 and 2008. The majority of these vessels primarily landed Nephrops or a combination of Hake, monkfish and megrim. Only eight vessels primarily landed whitefish (cod, haddock and whiting). A French decommissioning scheme was implemented in 2008 and 2009. A reduction in the French fleet operating in VIIe-k is expected as a result.

Table 7.15.1. Whiting in Divisions VIIe-k. Nominal Landings (t) as reported to ICES, and total landings as used by the Working Group.

|  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 130 | 158 | 160 | 107 | 112 | 159 | 295 | 317 | 304 | 111 | 145 | 228 | 205 | 268 |
| Denmark |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| France | 7,572 | 4,024 | 7,819 | 7,763 | 9,773 | 10,947 | 19,771 | 19,348 | 10,006 | 9,620 | 11,285 | 13,535 | 13,400 | 9,936 |
| Germany |  |  |  |  |  |  |  |  |  | 14 |  |  |  |  |
| Ireland | 1,511 | 1,227 | 2,241 | 1,309 | 1,518 | 2,036 | 1,651 | 1,764 | 1,403 | 1,875 | 3,630 | 5,053 | 6,077 | 6,115 |
| Netherlands |  | 398 |  | 124 |  |  |  |  |  |  |  |  |  | 8 |
| Spain |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 31 |
| UK (E/W/NI) | 1,192 | 986 | 751 | 910 | 1,098 | 1,632 | 1,326 | 1,829 | 2,023 | 1,393 | 1,776 | 1,624 | 1,803 | 1,724 |
| UK(Scotland) |  |  |  |  |  | 1 | 33 | 32 | 20 | 41 | 16 | 23 | 23 | 34 |
| United Kingdom |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Channel Islands |  |  | 2 | 2 | 2 |  |  |  |  |  |  |  | 1 | 1 |
| Total | 10,405 | 6,793 | 10,973 | 10,215 | 12,503 | 14,775 | 23,076 | 23,290 | 13,756 | 13,054 | 16,852 | 20,463 | 21,513 | 18,116 |
| Unallocated | 1,376 | 3,192 | -135 | -263 | 149 | 353 | -6,535 | -9,184 | -248 | -690 | -532 | -429 | 1,165 | 144 |
| Total as used by Working Group | 11,781 | 9,985 | 10,838 | 9,952 | 12,652 | 15,128 | 16,541 | 14,106 | 13,508 | 12,364 | 16,320 | 20,034 | 22,678 | 18,260 |


|  | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9 a}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Belgium | 449 | 479 | 448 | 194 | 171 | 149 | 129 | 180 | 218 | 128 | 127 | 122 | 87 |
| Denmark |  |  |  |  |  |  |  |  |  |  |  |  |  |
| France | 11,370 | 11,711 | $16,418^{\mathrm{b}}$ | 9,077 | 7,203 | 7,435 | 5,897 | 4,811 | 5,784 | 4,649 | 3,543 | 3,046 |  |
| Germany |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 6,893 | 5,226 | 5,807 | 4,795 | 5,008 | 5,332 | 4,093 | 4,215 | 5,709 | 4,521 | 4,764 | 2,330 | 2,328 |
| Netherlands |  | 1 |  |  | 5 | 4 | 9 | 18 | 60 | 40 | 64 | 23 | 29 |
| Spain | 24 | 53 | 21 | 11 | 9 | 12 | - | 76 | 56 | 70 | 21 | 8 |  |
| UK (E/W/NI) | 1,742 | 1,706 | 1,344 | 1,249 | 943 | 843 | 758 | 586 | 471 | 402 | 569 | 610 | 826 |
| UK(Scotland) | 42 | 68 | 3 | 2 | 11 | 12 | 5 | 7 | - | 6 | 4 | 7 |  |
| United Kingdom |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Channel Islands |  | 3 | 2 | 3 | 3 | 1 | 4 | 0 | 0 | 0 | 1 | 1 |  |
| Total | 20,520 | 19,247 | 24,043 | 15,331 | 13,353 | 13,788 | 10,895 | 9,893 | 12,298 | 9,816 | 9,093 | 6,147 | 3,270 |
| Unallocated | 12 | -2 | $-4,128$ | -466 | -583 | -642 | -312 | 61 | -269 | -283 | -146 | -410 | 2,439 |
| Total as used by |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Working Group | 20,532 | 19,245 | 19,915 | 14,865 | 12,770 | 13,146 | 10,583 | 9,954 | 12,030 | 9,533 | 8,948 | 5,737 | 5,708 |

[^16]Table 7.15.2. Whiting in Divisions VIIe-k. Raised length distributions for 2009 by country and fleet (Numbers in '000s).

| Length (cm) | France <br> VII fgh | UK (E+W) |  | Ireland |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Beam trawl VIIe-k | All gears (exc beam) VIIe-k | Scottish Seine VIIg | Otter trawl VIIg | Beam trawl VIIg | $\begin{gathered} \text { Gillnet } \\ \text { VIIg } \\ \hline \end{gathered}$ | Otter trawl VIIj | Scottish seine VIIj |
| 20 |  |  |  |  |  |  |  |  |  |
| 21 |  |  |  |  | 0.2 |  |  |  |  |
| 22 |  |  |  |  | 0.6 |  |  |  |  |
| 23 |  |  |  |  | 2.1 |  |  |  |  |
| 24 | 0.2 |  |  |  | 4.5 |  |  |  |  |
| 25 | 0.5 |  |  |  | 5.1 |  |  | 3.2 |  |
| 26 | 1.8 |  | 6.4 |  | 6.2 |  |  | 14.7 |  |
| 27 | 7.2 |  | 7.8 | 1.1 | 16.5 |  |  | 26.9 |  |
| 28 | 39.1 |  | 14.1 | 2.1 | 38.9 |  |  | 48.9 | 0.6 |
| 29 | 117.8 |  | 60.9 | 9.1 | 91.9 |  |  | 46.2 | 2.6 |
| 30 | 133.8 | 4.0 | 82.8 | 28.3 | 200.1 | 0.0 |  | 59.3 | 3.9 |
| 31 | 195.9 | 4.0 | 190.5 | 69.5 | 313.5 | 0.0 |  | 61.9 | 12.9 |
| 32 | 232.2 | 2.0 | 257.0 | 77.3 | 410.6 | 0.1 |  | 73.5 | 13.6 |
| 33 | 276.8 | 8.0 | 255.6 | 92.4 | 476.4 | 0.2 |  | 61.7 | 18.1 |
| 34 | 276.0 | 8.0 | 283.5 | 94.5 | 427.7 | 0.5 |  | 68.5 | 23.2 |
| 35 | 227.2 | 4.0 | 203.9 | 92.5 | 367.9 | 0.3 |  | 61.4 | 19.4 |
| 36 | 182.6 | 14.0 | 181.3 | 97.8 | 273.6 | 0.7 | 0.0 | 52.3 | 17.4 |
| 37 | 138.9 | 8.0 | 125.2 | 110.7 | 200.9 | 0.5 | 0.0 | 40.4 | 21.3 |
| 38 | 115.9 | 8.0 | 125.1 | 88.9 | 147.6 | 0.3 | 0.0 | 33.1 | 31.6 |
| 39 | 101.4 | 6.0 | 93.3 | 88.6 | 107.8 | 0.4 | 0.0 | 26.6 | 22.6 |
| 40 | 85.5 | 10.0 | 54.9 | 100.0 | 75.1 | 0.3 | 0.2 | 24.0 | 33.6 |
| 41 | 55.3 | 6.0 | 46.1 | 97.8 | 58.5 | 0.4 | 0.1 | 14.5 | 36.2 |
| 42 | 51.9 | 4.0 | 24.1 | 59.5 | 52.3 | 0.2 | 0.5 | 16.4 | 36.8 |
| 43 | 34.0 | 6.0 | 19.9 | 62.4 | 30.8 | 0.3 | 0.4 | 10.1 | 34.2 |
| 44 | 31.1 | 0.0 | 15.3 | 42.1 | 27.1 | 0.3 | 0.1 | 8.7 | 36.8 |
| 45 | 24.5 | 0.0 | 4.3 | 44.2 | 24.0 | 0.2 | 0.3 | 4.4 | 31.6 |
| 46 | 26.2 | 0.0 | 6.8 | 39.4 | 17.9 | 0.1 | 0.2 | 3.5 | 23.9 |
| 47 | 14.6 | 2.0 | 2.6 | 25.8 | 19.2 | 0.1 | 0.2 | 2.7 | 14.8 |
| 48 | 17.4 | 2.0 | 0.3 | 22.4 | 17.8 | 0.1 | 0.1 | 1.7 | 10.3 |
| 49 | 12.7 | 0.0 | 1.7 | 18.1 | 9.4 | 0.0 | 0.2 | 1.7 | 9.7 |
| 50 | 15.8 | 0.0 | 2.0 | 8.2 | 7.4 | 0.0 | 0.6 | 0.5 | 8.4 |
| 51 | 10.7 | 0.0 | 1.1 | 8.2 | 4.8 | 0.0 | 0.4 | 1.2 | 7.7 |
| 52 | 7.2 | 2.0 | 0.5 | 2.1 | 3.6 | 0.1 | 0.3 | 0.2 | 7.1 |
| 53 | 8.4 | 0.0 | 1.4 | 9.3 | 6.0 | 0.0 | 0.3 | 1.9 | 5.2 |
| 54 | 3.5 | 2.0 |  | 5.1 | 1.4 | 0.0 | 0.4 | 1.4 | 10.3 |
| 55 | 2.6 |  |  | 2.1 | 3.3 | 0.0 | 0.4 | 1.3 | 3.2 |
| 56 | 4.5 |  |  | 2.1 | 1.1 | 0.0 | 0.5 | 0.0 | 3.9 |
| 57 | 1.3 |  |  | 3.0 | 1.6 |  | 0.2 | 0.0 | 1.3 |
| 58 | 1.7 |  |  | 0.0 | 1.3 |  | 0.2 | 0.0 | 1.9 |
| 59 | 1.0 |  |  | 1.1 | 1.6 |  | 0.1 | 0.0 | 1.9 |
| 60 | 0.8 |  |  | 3.1 | 1.4 |  | 0.2 | 0.2 | 0.0 |
| 61 | 0.2 |  |  | 0.0 | 1.0 |  | 0.0 |  | 0.0 |
| 62 | 0.9 |  |  | 1.1 | 0.0 |  | 0.0 |  | 0.6 |
| 63 | 0.4 |  |  |  | 0.0 |  | 0.0 |  | 0.0 |
| 64 | 0.0 |  |  |  | 0.2 |  | 0.0 |  | 0.0 |
| 65 | 0.0 |  |  |  | 0.0 |  | 0.0 |  | 0.6 |
| 66 |  |  |  |  | 0.0 |  |  |  | 0.0 |
| 67 |  |  |  |  | 0.0 |  |  |  | 0.6 |
| 68 |  |  |  |  | 0.2 |  |  |  |  |
| 69 |  |  |  |  |  |  |  |  |  |
| Total N . | 2459.5 | 100.1 | 2068.6 | 1409.6 | 3459.1 | 5.2 | 6.1 | 773.1 | 508.1 |
| Total (t) | 1090.1 | 47.1 | 751.9 | 733.2 | 1296.0 | 2.7 | 6.6 | 283.2 | 351.6 |

Table 7.15.3. Whiting in Divisions VIIe-k. Landings numbers-at-age (‘000), examples of strong year classes are highlighted.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1982 | 0 | 2624 | 12523 | 9862 | 4564 | 880 | 41 | 23 |
| 1983 | 0 | 5867 | 9981 | 9059 | 3393 | 1319 | 195 | 10 |
| 1984 | 0 | 2854 | 18645 | 4697 | 1815 | 618 | 128 | 28 |
| 1985 | 0 | 3698 | 15538 | 8005 | 1380 | 289 | 96 | 33 |
| 1986 | 0 | 3769 | 15157 | 6465 | 2091 | 553 | 60 | 45 |
| 1987 | 0 | 5977 | 19376 | 8825 | 2467 | 587 | 112 | 60 |
| 1988 | 0 | 2315 | 26780 | 11400 | 1962 | 409 | 70 | 21 |
| 1989 | 0 | 602 | 17057 | 24243 | 3459 | 339 | 63 | 25 |
| 1990 | 0 | 3270 | 9249 | 19509 | 8654 | 749 | 62 | 21 |
| 1991 | 0 | 8339 | 11997 | 5578 | 11742 | 2700 | 143 | 3 |
| 1992 | 0 | 4964 | 20513 | 9198 | 1420 | 1275 | 435 | 39 |
| 1993 | 0 | 2304 | 22277 | 17939 | 2829 | 526 | 382 | 172 |
| 1994 | 0 | 1272 | 14110 | 25384 | 6165 | 1019 | 135 | 177 |
| 1995 | 0 | 540 | 15062 | 21854 | 14142 | 2242 | 310 | 92 |
| 1996 | 0 | 1345 | 7473 | 17783 | 12850 | 5486 | 775 | 114 |
| 1997 | 0 | 609 | 4451 | 11734 | 21209 | 7322 | 2787 | 720 |
| 1998 | 0 | 1182 | 6680 | 10938 | 12758 | 13240 | 2865 | 882 |
| 1999 | 0 | 4163 | 10223 | 12444 | 8406 | 8733 | 6479 | 1188 |
| 2000 | 0 | 3575 | 9357 | 10328 | 5468 | 2351 | 1993 | 1845 |
| 2001 | 0 | 336 | 11648 | 11076 | 5135 | 2061 | 745 | 275 |
| 2002 | 0 | 1067 | 5962 | 19658 | 5732 | 1064 | 274 | 63 |
| 2003 | 0 | 462 | 3599 | 8264 | 11530 | 1675 | 264 | 20 |
| 2004 | 0 | 1209 | 4141 | 5963 | 6755 | 5978 | 496 | 69 |
| 2005 | 0 | 768 | 6169 | 8141 | 5008 | 4551 | 3456 | 147 |
| 2006 | 0 | 1366 | 6342 | 7631 | 3672 | 1767 | 1148 | 581 |
| 2007 | 0 | 988 | 5598 | 8479 | 4984 | 1535 | 412 | 226 |
| 2008 | 0 | 1269 | 3710 | 5948 | 2923 | 700 | 173 | 31 |
| 2009 | 0 | 341 | 4194 | 5693 | 2768 | 695 | 165 | 36 |

Table 7.15.4. Whiting in Divisions VIIe-k. Landings weights-at-age (kg).

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1982 | 0.000 | 0.245 | 0.279 | 0.395 | 0.557 | 0.646 | 1.193 | 1.593 |
| 1983 | 0.000 | 0.273 | 0.328 | 0.441 | 0.545 | 0.678 | 0.731 | 1.652 |
| 1984 | 0.000 | 0.227 | 0.286 | 0.457 | 0.656 | 0.807 | 1.060 | 1.514 |
| 1985 | 0.000 | 0.233 | 0.335 | 0.433 | 0.631 | 1.008 | 1.157 | 0.980 |
| 1986 | 0.000 | 0.198 | 0.277 | 0.493 | 0.585 | 0.781 | 1.469 | 1.680 |
| 1987 | 0.000 | 0.222 | 0.284 | 0.398 | 0.658 | 0.877 | 0.897 | 0.990 |
| 1988 | 0.000 | 0.224 | 0.303 | 0.416 | 0.628 | 0.977 | 1.322 | 1.374 |
| 1989 | 0.000 | 0.201 | 0.281 | 0.376 | 0.593 | 0.980 | 1.444 | 1.877 |
| 1990 | 0.000 | 0.226 | 0.260 | 0.328 | 0.452 | 0.722 | 1.083 | 1.721 |
| 1991 | 0.000 | 0.220 | 0.291 | 0.355 | 0.395 | 0.534 | 0.834 | 1.695 |
| 1992 | 0.000 | 0.208 | 0.289 | 0.388 | 0.472 | 0.623 | 0.739 | 1.084 |
| 1993 | 0.086 | 0.205 | 0.286 | 0.379 | 0.589 | 0.831 | 0.963 | 1.360 |
| 1994 | 0.000 | 0.249 | 0.300 | 0.404 | 0.637 | 0.915 | 0.982 | 1.222 |
| 1995 | 0.090 | 0.202 | 0.275 | 0.382 | 0.527 | 0.844 | 1.124 | 1.197 |
| 1996 | 0.000 | 0.229 | 0.266 | 0.346 | 0.460 | 0.598 | 0.616 | 1.058 |
| 1997 | 0.000 | 0.196 | 0.277 | 0.329 | 0.406 | 0.536 | 0.714 | 1.005 |
| 1998 | 0.000 | 0.188 | 0.270 | 0.333 | 0.396 | 0.452 | 0.567 | 0.896 |
| 1999 | 0.000 | 0.222 | 0.298 | 0.352 | 0.426 | 0.441 | 0.497 | 0.633 |
| 2000 | 0.101 | 0.250 | 0.326 | 0.419 | 0.510 | 0.573 | 0.585 | 0.597 |
| 2001 | 0.000 | 0.265 | 0.286 | 0.393 | 0.521 | 0.624 | 0.761 | 0.820 |
| 2002 | 0.082 | 0.217 | 0.293 | 0.363 | 0.519 | 0.682 | 0.810 | 1.022 |
| 2003 | 0.000 | 0.211 | 0.281 | 0.369 | 0.447 | 0.603 | 0.831 | 1.149 |
| 2004 | 0.086 | 0.218 | 0.303 | 0.376 | 0.433 | 0.492 | 0.523 | 0.754 |
| 2005 | 0.101 | 0.246 | 0.318 | 0.396 | 0.506 | 0.509 | 0.487 | 0.595 |
| 2006 | 0.112 | 0.232 | 0.299 | 0.414 | 0.545 | 0.585 | 0.586 | 0.707 |
| 2007 | 0.000 | 0.206 | 0.290 | 0.389 | 0.492 | 0.603 | 0.564 | 0.673 |
| 2008 | 0.116 | 0.235 | 0.291 | 0.378 | 0.512 | 0.617 | 0.754 | 1.124 |
| 2009 | 0.000 | 0.245 | 0.322 | 0.405 | 0.504 | 0.592 | 0.669 | 0.902 |

Table 7.15.5. Whiting in Divisions VIIe-k. Stock weights-at-age (kg).

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 0.157 | 0.270 | 0.345 | 0.474 | 0.607 | 0.843 | 1.403 | 1.255 | 0.688 | 0.688 |
| 1983 | 0 | 0.167 | 0.276 | 0.363 | 0.498 | 0.632 | 0.826 | 1.313 | 1.256 | 0.732 | 0.732 |
| 1984 | 0 | 0.192 | 0.282 | 0.371 | 0.521 | 0.709 | 0.847 | 1.188 | 1.270 | 0.723 | 0.723 |
| 1985 | 0 | 0.179 | 0.272 | 0.389 | 0.534 | 0.738 | 1.030 | 1.187 | 1.382 | 1.046 | 0.957 |
| 1986 | 0 | 0.183 | 0.259 | 0.370 | 0.543 | 0.756 | 1.020 | 1.223 | 1.513 | 1.145 | 0.98 |
| 1987 | 0 | 0.171 | 0.253 | 0.367 | 0.533 | 0.752 | 1.059 | 1.261 | 1.474 | 1.585 | 0.864 |
| 1988 | 0 | 0.186 | 0.252 | 0.342 | 0.531 | 0.784 | 1.050 | 1.322 | 1.685 | 1.465 | 0.768 |
| 1989 | 0 | 0.173 | 0.249 | 0.331 | 0.477 | 0.760 | 1.114 | 1.439 | 1.643 | 1.853 | 0.599 |
| 1990 | 0 | 0.166 | 0.247 | 0.317 | 0.427 | 0.651 | 1.007 | 1.524 | 1.461 | 1.465 | 0.842 |
| 1991 | 0 | 0.151 | 0.248 | 0.317 | 0.396 | 0.553 | 0.815 | 1.310 | 1.154 | 1.032 | 0.929 |
| 1992 | 0 | 0.174 | 0.253 | 0.327 | 0.421 | 0.551 | 0.736 | 1.133 | 1.105 | 0.866 | 1.216 |
| 1993 | 0 | 0.166 | 0.251 | 0.340 | 0.470 | 0.637 | 0.779 | 1.034 | 1.337 | 0.954 | 1.126 |
| 1994 | 0 | 0.175 | 0.254 | 0.340 | 0.487 | 0.715 | 0.906 | 1.077 | 1.258 | 1.405 | 1.158 |
| 1995 | 0 | 0.108 | 0.259 | 0.346 | 0.476 | 0.711 | 0.861 | 0.994 | 1.047 | 1.341 | 1.044 |
| 1996 | 0 | 0.135 | 0.256 | 0.328 | 0.430 | 0.626 | 0.820 | 0.942 | 0.990 | 1.107 | 1.035 |
| 1997 | 0 | 0.110 | 0.245 | 0.307 | 0.396 | 0.525 | 0.645 | 0.830 | 1.123 | 0.912 | 0.912 |
| 1998 | 0 | 0.148 | 0.238 | 0.293 | 0.378 | 0.453 | 0.585 | 0.747 | 1.043 | 0.968 | 0.968 |
| 1999 | 0 | 0.112 | 0.245 | 0.324 | 0.419 | 0.491 | 0.518 | 0.677 | 0.779 | 0.725 | 0.725 |
| 2000 | 0 | 0.144 | 0.253 | 0.357 | 0.465 | 0.556 | 0.611 | 0.711 | 0.685 | 0.895 | 0.895 |
| 2001 | 0 | 0.182 | 0.259 | 0.370 | 0.490 | 0.612 | 0.676 | 0.802 | 0.649 | 0.995 | 0.995 |
| 2002 | 0 | 0.193 | 0.248 | 0.361 | 0.480 | 0.627 | 0.795 | 1.009 | 0.850 | 1.062 | 1.062 |
| 2003 | 0 | 0.187 | 0.244 | 0.332 | 0.439 | 0.560 | 0.693 | 0.886 | 1.202 | 0.875 | 1.127 |
| 2004 | 0 | 0.167 | 0.253 | 0.333 | 0.449 | 0.541 | 0.652 | 0.892 | 1.380 | 1.38 | 1.38 |
| 2005 | 0 | 0.163 | 0.256 | 0.346 | 0.484 | 0.535 | 0.582 | 0.765 | 1.431 | 1.431 | 1.431 |
| 2006 | 0 | 0.177 | 0.280 | 0.390 | 0.553 | 0.624 | 0.647 | 0.832 | 0.990 | 0.799 | 0.799 |
| 2007 | 0 | 0.204 | 0.285 | 0.403 | 0.566 | 0.666 | 0.727 | 0.951 | 0.811 | 0.633 | 0.633 |
| 2008 | 0 | 0.227 | 0.298 | 0.397 | 0.549 | 0.659 | 0.714 | 0.920 | 0.527 | 0.467 | 0.467 |
| 2009 | 0 | 0.220 | 0.286 | 0.380 | 0.525 | 0.631 | 0.723 | 0.981 | 0.540 | 0.54 | 0.54 |

Table 7.15.6. Whiting in Divisions VIIe-k. Summary of discard data in 2009 provided the Working Group.

|  |  |  |  | Sampling |  |  |  |  |  |  | Discard Rates |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Year | Quarter | Gear/Fleet | Trips | Hauls | Numbers Retained | Weight Retained | Number Discarded | Weight Discarded | Units | Number | Weight |
| France | 2009 | Q1-3 | OT VIIe-k Crustacean | 22 | 173 | 2555 | 880 | 4569 | 640 | No. \& KG Sampled | 64\% | 42\% |
| France | 2009 | Q2-4 | OT VIIe-k Demersal fish | 42 | 357 | 1396 | 666 | 1842 | 286 | No. \& KG Sampled | 57\% | 30\% |
| UK | 2009 | 1 | All Gears | 33 | 453 | 3175 | 1409 | 5612 | 1026 | Raised No. \& KG Sampled | 64\% | 42\% |
| UK | 2009 | 2 | All Gears | 22 | 164 | 2387 | 841 | 2497 | 560 | Raised No. \& KG Sampled | 51\% | 40\% |
| UK | 2009 | 3 | All Gears | 29 | 398 | 2377 | 796 | 2994 | 644 | Raised No. \& KG Sampled | 56\% | 45\% |
| UK | 2009 | 4 | All Gears | 12 | 73 | 3294 | 1629 | 2088 | 315 | Raised No. \& KG Sampled | 39\% | 16\% |
| UK | 2009 | All | All Gears | 96 | 1088 | 11233 | 4674 | 13191 | 2546 | Raised No. \& KG Sampled | 54\% | 35\% |
| Ireland | 2009 | All | Otter Trawls VIIg | 16 | 175 | 5703 | 1969 | 18909 | 3051 | No. '000s \& tonnes raised to Fleet | 77\% | 61\% |
| Ireland | 2009 | All | Otter Trawls VIIj | 15 | 200 | 106 | 43 | 367 | 58 | No. '000s \& tonnes raised to Fleet | 78\% | 57\% |

Table 7.15.7. Whiting in Divisions VIIe-k. Standardised survey abundance indices of age groups 0-3.

| Survey Units | UK-WCGFS <br> No. per min |  |  | UK-BCCSBTS-S No. per km towed |  | FR-EVHOE <br> No. per 30 min haul |  |  |  | IR-GFS-7g\&j <br> No. per 30 min haul |  |  |  | IR-GFS-7g-Swept Area No. per 10 kmsq |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1-gp | 2-gp | 3-gp | 0-gp | 1-gp | 0-gp | 1-gp | 2-gp | 3-gp | 0-gp | 1-gp | 2-gp | 3-gp | 0-gp | 1-gp | 2-gp | 3-gp |
| 1987 | 0.36 | 1.61 | 0.16 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1988 | 0.24 | 0.23 | 0.06 | 0.1 | 0.9 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989 | 0.25 | 0.73 | 0.49 | 0.9 | 1.1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1990 | 0.02 | 0.06 | 0.25 | 5.2 | 0.5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 | 0.21 | 0.01 | 0.01 | 4.4 | 1.4 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 1.31 | 0.53 | 0.11 | 6.7 | 1.3 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 | 4.88 | 0.92 | 0.27 | 10.0 | 1.7 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 | 8.99 | 1.33 | 0.92 | 2.7 | 1.5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995 | 0.59 | 5.52 | 1.43 | 2.3 | 1.5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1996 | 0.52 | 1.51 | 1.39 | 4.6 | 1.5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 | 0.73 | 0.56 | 0.18 | 10.7 | 0.5 | 31 | 24 | 9 | 8.5 |  |  |  |  |  |  |  |  |
| 1998 | 1.19 | 0.77 | 0.53 | 5.3 | 0.5 | 48 | 15 | 7.9 | 1.2 |  |  |  |  |  |  |  |  |
| 1999 | 0.84 | 0.50 | 0.15 | 15.1 | 1.0 | 261 | 62 | 18 | 5.1 |  |  |  |  | 24175 | 7307 | 1881 | 633 |
| 2000 | 14.91 | 0.93 | 0.29 | 1.2 | 3.1 | 31 | 77 | 23 | 2.9 |  |  |  |  | 6077 | 15835 | 3116 | 190 |
| 2001 | 2.49 | 1.35 | 0.24 | 1.7 | 0.5 | 23 | 35 | 49 | 8 |  |  |  |  | 4650 | 2836 | 13871 | 1849 |
| 2002 | 3.35 | 1.80 | 3.04 | 5.3 | 0.3 | 39 | 15 | 11 | 10 |  |  |  |  | 2468 | 3664 | 1719 | 1252 |
| 2003 | 3.20 | 2.51 | 2.48 | 3.9 | 0.1 | 47 | 58 | 27 | 20 | 127 | 88 | 38 | 11 | 6061 | 2219 | 1027 | 413 |
| 2004 | 2.00 | 1.80 | 0.99 | 10.3 | 0.1 | 28 | 108 | 31 | 14 | 295 | 95 | 48 | 10 | 9778 | 3444 | 655 | 321 |
| 2005 | Survey discontinued |  |  | 6.4 | 0.0 | 44 | 16 | 5 | 2 | 83 | 106 | 29 | 10 | 1146 | 3177 | 1573 | 422 |
| 2006 |  |  |  | 4.3 | 0.3 | 15 | 10 | 3 | 1 | 373 | 161 | 50 | 10 | 15260 | 5883 | 2175 | 707 |
| 2007 |  |  |  | 7.7 | 0.7 | 178 | 46 | 4 | 1 | 332 | 218 | 47 | 7 | 9951 | 8081 | 2718 | 455 |
| 2008 |  |  |  | 25.1 | 0.7 | 365 | 45 | 10 | 3 | 402 | 140 | 44 | 11 | 16344 | 5554 | 2238 | 475 |
| 2009 |  |  |  | 6.7 | 0.6 | 30 | 68 | 31 | 6 | 346 | 289 | 65 | 17 | 11053 | 10819 | 2154 | 589 |

Table 7.15.8. Whiting in Divisions VIIe-k. Available commercial and survey tuning-series, ages and years used in the assessment are highlighted in bold.

Whiting in the Celtic Sea VIIe-k Tuning data WGCSE 2009 (Sarah Davie 27/04/09) 114
FR-GADOID-Early: French Gadoid trawlers (FU5) - Effort, No. of whiting/age/1000 hours fished, Year, Live weight (t)


FR-GADOID-late: French Gadoid trawlers (FU5) - Effort, No. of whiting/age/1000 hours fished, Year, Live weight (t)

| 1993 | 2008 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | $\bigcirc$ | 1 |  |  |  |  |  |  |  |  |  |  |
| 1 | 11 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 4736 | 57675 | 35630 | 5286 | 825 | 883 | 469 | 40 | 20 | 6 | 0 | \#1993 | 7815t |
| 1000 | 448 | 26922 | 65786 | 18395 | 2948 | 289 | 454 | 125 | 80 | 0 | 0 | \#1994 | 9236t |
| 1000 | 86 | 10737 | 43840 | 34895 | 7662 | 1360 | 248 | 0 | 28 | 32 | 0 | \#1995 | 9186t |
| 1000 | 8 | 2509 | 34872 | 31293 | 13650 | 1708 | 328 | 32 | 31 | 29 | 0 | \#1996 | 6028t |
| 1000 | 0 | 3641 | 17743 | 45915 | 14168 | 4338 | 721 | 63 | 12 | 0 | 0 | \#1997 | 7218t |
| 1000 | 3827 | 17367 | 32394 | 25399 | 30762 | 21832 | 3285 | 631 | 186 | 0 | 0 | \#1998 | 7674t |
| 1000 | 3457 | 15689 | 29265 | 22945 | 27790 | 19723 | 2967 | 570 | 168 | 0 | 0 | \#1999 | 9102t |
| 1000 | 4987 | 23934 | 29232 | 15124 | 6851 | 7110 | 5976 | 1306 | 132 | 10 | 0 | \#2000 | 6053t |
| 1000 | 213 | 23745 | 25724 | 9253 | 3440 | 1465 | 593 | 539 | 114 | 57 | 0 | \#2001 | 4624t |
| 1000 | 405 | 9574 | 48049 | 13052 | 2399 | 816 | 136 | 59 | 27 | 25 | 0 | \#2002 | 4799t |
| 1000 | 13 | 2004 | 15027 | 33581 | 3776 | 542 | 94 | 48 | 67 | 13 | 3 | \#2003 | 2975t |
| 1000 | 238 | 4747 | 10190 | 18892 | 20570 | 1688 | 269 | 17 | 0 | 0 | 0 | \#2004 | 2589t |
| 1000 | 278 | 11772 | 23815 | 15806 | 17601 | 15832 | 418 | 54 | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | \#2005 | 3659t |
| 1000 | 295 | 16943 | 35200 | 15517 | 7869 | 5396 | 2180 | 142 |  | 0 | 0 | \#2006 | 2795t |
| 1000 | 369 | 13147 | 23994 | 12964 | 2496 | 461 | 400 | 460 | 53 | $\bigcirc$ | 0 | \#2007 | 1898t |
| 1000 | 257 | 8841 | 14651 | 10665 | 2942 | 586 | 50 | 65 | 0 | 0 | 0 | \#2008 | 1133t |

FR-NEPHROPS-Early: French Nephrops trawlers (FU8) - Effort, No. whiting/age/1000 hours fished, Year, Live weight ( t )

| 1987 | 1992 |  |  |  |  |
| :--- | :--- | :---: | :--- | :--- | :--- |
| 1 | 1 | 0 | 1 |  |  |
| 1 | 11 |  |  |  |  |
| 1000 | 917 | 3681 | 2247 | 761 | 176 |
| 1000 | 632 | 7960 | 3610 | 918 | 165 |
| 1000 | 131 | 4874 | 6866 | 1294 | 128 |
| 1000 | 321 | 1139 | 3596 | 2297 | 279 |
| 1000 | 1048 | 2312 | 982 | 1745 | 498 |
| 1000 | 1542 | 6078 | 3348 | 478 | 571 |

FR-NEPHROPS-Late: French Nephrops trawlers (FU8) - Effort, No. whiting/age/1000 hours fished, Year, Live weight (t)

| 1993 | 2008 |  | 1 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 11 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 766 | 6928 | 5695 | 1001 | 163 | 86 | 74 | 1 | 2 | 0 | $\bigcirc$ | \#1993 | 1356 t |
| 1000 | 184 | 6145 | 8313 | 1840 | 214 | 17 | 16 | 5 | 2 | 0 | 0 | \#1994 | 1565t |
| 1000 | 29 | 2217 | 7580 | 4802 | 697 | 91 | 20 | $\bigcirc$ | 3 | 3 | 0 | \#1995 | 1446t |
| 1000 | 2 | 979 | 5599 | 4992 | 2359 | 305 | 55 | 4 | 1 | 7 | 0 | \#1996 | 1230t |
| 1000 | 0 | 737 | 3511 | 10406 | 4124 | 1231 | 275 | 23 | 1 | 0 | 0 | \#1997 | 1393t |
| 1000 | 58 | 1042 | 2567 | 4299 | 5925 | 1236 | 239 | 46 | 2 | 0 | 0 | \#1998 | 881t |
| 1000 | 1253 | 4408 | 4764 | 3762 | 3867 | 3563 | 575 | 136 | 8 | 0 | $\bigcirc$ | \#1999 | 1190t |
| 1000 | 277 | 2381 | 3085 | 2213 | 923 | 836 | 959 | 232 | 23 | 0 | 0 | \#2000 | 869t |
| 1000 | 104 | 2948 | 3131 | 1531 | 557 | 213 | 106 | 95 | 36 | 8 | 0 | \#2001 | 548t |
| 1000 | 27 | 747 | 4007 | 1455 | 462 | 170 | 69 | 13 | 14 | 7 | 0 | \#2002 | 550 t |


| 1000 | 5 | 311 | 1708 | 3944 | 574 | 95 | 27 | 7 | 1 | 0 | 2 | $\# 2003$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $543 t$ |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 47 | 748 | 1090 | 2045 | 2726 | 233 | 49 | 6 | 0 | 0 | 0 | $\# 2004$ |
| $435 t$ |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 104 | 1285 | 1926 | 1133 | 1266 | 1283 | 54 | 2 | 0 | 0 | 0 | $\# 2005$ |
| $378 t$ |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 46 | 802 | 1299 | 591 | 299 | 187 | 101 | 12 | 0 | 0 | 0 | $\# 2006$ |
| 1000 | 138 | 981 | 1159 | 604 | 137 | 26 | 19 | 16 | 5 | 0 | 0 | $\# 2007$ |
| $106 t$ |  |  |  |  |  |  |  |  |  |  |  |  |
| 1000 | 41 | 506 | 565 | 408 | 96 | 19 | 7 | 2 | 0 | 0 | 0 | $\# 2008$ |

FR-EVHOE: Thalassa Survey - No. whiting at age/30 min, Year

## 19972009

| 1 | 1 | 0.75 | 1 |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 8 |  |  |  |  |  |  |  |  |  |
| 1 | 30.82 | 23.85 | 8.93 | 8.47 | 10.38 | 1.93 | 0.24 | 0.00 | 0.00 | \#1997 |
| 1 | 48.10 | 15.15 | 7.88 | 1.23 | 1.67 | 0.55 | 0.18 | 0.02 | 0.00 | \#1998 |
| 1 | 260.66 | 62.15 | 17.64 | 5.09 | 1.92 | 1.67 | 1.18 | 0.15 | 0.13 | \#1999 |
| 1 | 30.62 | 76.50 | 23.18 | 2.85 | 1.17 | 0.33 | 0.18 | 0.50 | 0.06 | \#2000 |
| 1 | 22.77 | 35.46 | 48.80 | 8.12 | 0.79 | 0.14 | 0.11 | 0.02 | 0.04 | \#2001 |
| 1 | 38.50 | 15.33 | 11.00 | 9.58 | 0.82 | 0.00 | 0.00 | 0.00 | 0.00 | \#2002 |
| 1 | 46.62 | 58.30 | 27.11 | 19.94 | 14.74 | 0.05 | 0.01 | 0.00 | 0.00 | \#2003 |
| 1 | 28.23 | 108.11 | 31.11 | 14.36 | 6.98 | 3.98 | 0.00 | 0.00 | 0.00 | \#2004 |
| 1 | 44.14 | 15.85 | 5.19 | 1.89 | 1.15 | 0.63 | 0.16 | 0.00 | 0.00 | \#2005 |
| 1 | 14.60 | 9.53 | 3.45 | 1.18 | 0.30 | 0.03 | 0.00 | 0.01 | 0.00 | \#2006 |
| 1 | 178.39 | 46.30 | 4.34 | 0.68 | 0.36 | 0.07 | 0.00 | 0.00 | 0.01 | \#2007 |
| 1 | 364.99 | 44.55 | 10.17 | 3.27 | 1.43 | 0.14 | 0.00 | 0.00 | 0.03 | \#2008 |
| 1 | 29.93 | 68.10 | 30.54 | 6.47 | 1.34 | 0.02 | 0.01 | 0.00 | 0.00 | \#2009 |

UK-WCGFS : UK (E+W) PHHT Groundfish Survey in VIIf\&g - Effort mins towed, no.s at age, Year, Vessel (final survey in 2004)

| 1987 | 2004 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 1 | 0.15 | 0.25 |  |  |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |  |  |
| 360 | 129 | 580 | 57 | 8 | 6 | 0 | 0 | \#1987 | Cirolana |
| 540 | 129 | 125 | 31 | 3 | 3 | 0 | Cirolana |  |  |
| 540 | 137 | 393 | 267 | 21 | 4 | 2 | 0 | \#1989 | Cirolana |
| 540 | 11 | 31 | 137 | 55 | 9 | 1 | 0 | \#1990 | Cirolana |
| 482 | 99 | 6 | 3 | 11 | 9 | 1 | 0 | \#1991 | Cirolana |
| 840 | 1097 | 441 | 94 | 28 | 22 | 6 | 1 | \#1992 | Cirolana |
| 840 | 4101 | 772 | 229 | 29 | 4 | 8 | 3 | \#1993 | Cirolana |
| 535 | 4809 | 713 | 490 | 70 | 17 | 1 | 3 | \#1994 | Cirolana |
| 1320 | 777.4 | 7282.9 | 1891.2 | 595 | 82.2 | 18.6 | 11.3 | \#1995 | Cirolana |
| 1475 | 773 | 2225 | 2050 | 391 | 148 | 11 | 2 | \#1996 | Corystes |
| 1519 | 1113 | 852 | 280 | 646 | 226 | 60 | 5 | \#1997 | Cirolana |
| 900 | 1071.5 | 691.5 | 477 | 343.3 | 104.8 | 13.3 | 12.5 | \#1998 | Cirolana |
| 900 | 760.2 | 453.9 | 139.4 | 52.1 | 47.8 | 90.2 | 30.5 | \#1999 | Cirolana |
| 1038 | 15471.8 | 962.8 | 296.4 | 118.9 | 47.2 | 51 | 50.6 | \#2000 | Cirolana |
| 880 | 2195.3 | 1186.5 | 206.8 | 35.4 | 2 | 7.6 | 1 | \#2001 | Cirolana |
| 762 | 2551.5 | 1368.9 | 2313.6 | 155.9 | 75.7 | 1.2 | 4.4 | \#2002 | Cirolana |
| 863 | 2765.7 | 2169.9 | 2138.8 | 1665.8 | 157.9 | 0 | 0 | \#2003 | Cirolana |
| 860 | 1716.8 | 1548.2 | 852.1 | 203.6 | 184.3 | 2 | 0 | \#2004 | Cefas Endeavour |

UK BT SURVEY : (Sept) - Prime stations only (VIIf) Effort (km towed), numbers at age per Km towed

| 1988 | 2009 |  |  |
| :--- | :--- | :--- | :--- |
| 1 | 1 | 0.75 | 0.85 |
| 0 | 1 |  |  |
| 74.12 | 6 | 66 | $\# 1988$ |
| 91.91 | 80 | 104 | $\# 1989$ |
| 69.86 | 363 | 37 | $\# 1990$ |
| 123.41 | 540 | 175 | $\# 1991$ |
| 125.08 | 839 | 164 | $\# 1992$ |
| 127.67 | 1279 | 213 | $\# 1993$ |
| 120.82 | 330 | 182 | $\# 1994$ |
| 104.14 | 240 | 154 | $\# 1995$ |
| 122.11 | 557 | 188 | $\# 1996$ |
| 115.63 | 1238 | 56 | $\# 1997$ |
| 104.7 | 553 | 49 | $\# 1998$ |
| 117.11 | 1770 | 116 | $\# 1999$ |
| 105.99 | 128 | 333 | $\# 2009$ |
| 118.22 | 204 | 56 | $\# 2001$ |
| 113.03 | 602 | 36 | $\# 2002$ |
| 111.92 | 442 | 6 | $\# 2003$ |
| 101.92 | 1053 | 6 | $\# 2004$ |
| 119.11 | 760 | 5 | $\# 2005$ |
| 120.56 | 520 | 31 | $\# 2006$ |
| 118.59 | 910 | 81 | $\# 2007$ |

Tows 15 minute duration - raised here to 30 minutes Tows 15 minute duration - raised here to 30 minutes


IR-7G\&J-OT : Irish Otter Trawl Fleet (Areas VIIg\&j) - Effort in hours, no.s @ age, Year, Live weight (t), LPUE (kg/h)

| 1995 | 2009 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |
| 1 | 4 |  |  |  |  |  |  |  |  |
| 157085 | 679 | 2281 | 1889 | 1333 | \# |  |  | \#1995 |  |
| 130257 | 164 | 1549 | 1889 | 905 | \# |  |  | \#1996 |  |
| 148276 | 170 | 756 | 1488 | 1247 | \# |  |  | \#1997 |  |
| 161909 | 180 | 933 | 980 | 736 | \# |  |  | \#1998 |  |
| 92195 | 388 | 960 | 962 | 449 | \# |  |  | \#1999 |  |
| 125229 | 619 | 1042 | 808 | 500 | \#228 | 103 | 65 | \#2000 | 1506.6t 12.03 |
| 137086 | 91 | 2224 | 1538 | 1046 | \#412 | 125 | 48 | \#2001 | 2227.9 t 16.25 |
| 168134 | 291 | 1140 | 2615 | 613 | \#86 | 13 | 6 | \#2002 | 1761.4 t 10.48 |
| 198059 | 147 | 878 | 1640 | 1195 | \#155 | 8 | 0 | \#2003 | 1544.6t 7.80 |
| 188948 | 132 | 628 | 1763 | 1002 | \#428 | 42 | 2 | \#2004 | 2243.9 t 11.88 |
| 198315 | 96 | 1743 | 2848 | 1226 | \#1162 | 745 | 31 | \#2005 | 3730.4 t 18.81 |
| 185083 | 188 | 1900 | 2070 | 950 | \#427 | 283 | 127 | \#2006 | 3008.2 t 16.25 |
| 217009 | 78 | 1063 | 3112 | 2305 | \#614 | 141 | 70 | \#2007 | 3597.2 t 16.58 |
| 192317 | 131 | 860 | 1038 | 677 | \#173 | 55 | 7 | \#2008 | 1269.3t 6.60 |
| 207631 | 216 | 894 | 1471 | 675 | \#283 | 69 | 17 | \#2009 | 1573.3t 7.58 |

IR-GFS-7G\&J : Irish Groundfish Survey in VIIg\&j (IBTS 4th Qtr) - Whiting no. @ age (Interim indices: New Celtic Explorer series)

| 2003 | 2009 |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 1 | 0.79 | 0.92 |  |  |  |  |  |
| 0 | 6 |  |  |  |  |  |  |  |
| 1612 | 6836 | 4714 | 2064 | 582 | 96 | 12 | 0 | $\# 2003$ |
| 1700 | 16710 | 5405 | 2733 | 570 | 170 | 115 | 10 | $\# 2004$ |
| 1726 | 4761 | 6085 | 1655 | 573 | 142 | 75 | 101 | $\# 2005$ |
| 1947 | 24194 | 10418 | 3250 | 637 | 100 | 3 | 25 | $\# 2006$ |
| 2042 | 22609 | 14869 | 3182 | 508 | 82 | 39 | 10 | $\# 2007$ |
| 2012 | 26990 | 9362 | 2957 | 734 | 135 | 6 | 8 | $\# 2008$ |
| 1765 | 20379 | 17026 | 3845 | 989 | 196 | 41 | 0 | $\# 2009$ |

Table 7.15.9. Whiting in Divisions VIIe-k. Landings (t), lpue of French and Irish fleets, and Effort ('000 h) of French, Irish and UK fleets.

|  | FR-Gadoid VII fg French gadoid trawlers |  |  | FR-Nephrops VII fg French <br> Nephrops trawlers |  |  | IR-OTB-7G <br> Irish otter trawlers <br> VIIg |  |  | IR-OTB-7J <br> Irish otter trawlers <br> VIIj |  |  | $\begin{aligned} & \text { UK (E\&W) in VIIe-k } \\ & \text { Beam Otter } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Landings | Effort ${ }^{4}$ | LPUE $^{3}$ | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Effort ${ }^{4}$ | Effort ${ }^{4}$ |
| 1983 | 5,742 | 109 | 53 | 470 | 207 | 2 |  |  |  |  |  |  | 135 | 82 |
| 1984 | 4,598 | 84 | 55 | 340 | 173 | 2 |  |  |  |  |  |  | 131 | 87 |
| 1985 | 4,514 | 89 | 51 | 651 | 185 | 4 |  |  |  |  |  |  | 152 | 90 |
| 1986 | 5,049 | 116 | 44 | 374 | 146 | 3 |  |  |  |  |  |  | 136 | 85 |
| 1987 | 6,859 | 137 | 50 | 588 | 177 | 3 |  |  |  |  |  |  | 177 | 84 |
| 1988 | 7,921 | 200 | 40 | 844 | 156 | 5 |  |  |  |  |  |  | 195 | 89 |
| 1989 | 8,974 | 231 | 39 | 891 | 159 | 6 |  |  |  |  |  |  | 198 | 84 |
| 1990 | 7,897 | 188 | 42 | 671 | 196 | 3 |  |  |  |  |  |  | 208 | 99 |
| 1991 | 7,525 | 167 | 45 | 527 | 187 | 3 |  |  |  |  |  |  | 203 | 77 |
| 1992 | 6,460 | 173 | 37 | 1,153 | 234 | 5 |  |  |  |  |  |  | 196 | 86 |
| 1993 | 7,815 | 201 | 39 | 1,356 | 223 | 6 |  |  |  |  |  |  | 208 | 62 |
| 1994 | 9,236 | 171 | 54 | 1,565 | 223 | 7 |  |  |  |  |  |  | 220 | 54 |
| 1995 | 9,186 | 171 | 54 | 1,446 | 202 | 7 | 829 | 64 | 13 | 1,305 | 94 | 14 | 243 | 52 |
| 1996 | 6,028 | 152 | 40 | 1,230 | 179 | 7 | 906 | 60 | 15 | 803 | 70 | 11 | 261 | 61 |
| 1997 | 7,218 | 195 | 37 | 1,393 | 149 | 9 | 1,066 | 65 | 16 | 783 | 83 | 9 | 265 | 67 |
| 1998 | 9,102 | 172 | 53 | 881 | 125 | 7 | 813 | 72 | 11 | 545 | 90 | 6 | 255 | 62 |
| 1999 | 9,102 | 191 | 48 | 1,190 | 130 | 9 | 946 | 52 | 18 | 247 | 41 | 6 | 251 | 98 |
| 2000 | 6,053 | 157 | 38 | 869 | 161 | 5 | 990 | 61 | 16 | 517 | 65 | 8 | 259 | 104 |
| 2001 | 4,624 | 174 | 27 | 548 | 137 | 4 | 1,286 | 69 | 19 | 942 | 68 | 14 | 273 | 85 |
| 2002 | 4,841 | 165 | 29 | 550 | 142 | 4 | 1,004 | 78 | 13 | 758 | 90 | 8 | 249 | 83 |
| 2003 | 2,975 | 125 | 24 | 543 | 161 | 3 | 1,051 | 87 | 12 | 494 | 111 | 4 | 282 | 72 |
| 2004 | 2,589 | 107 | 24 | 435 | 127 | 3 | 1,932 | 97 | 20 | 312 | 92 | 3 | 274 | 76 |
| 2005 | 3,787 | 93 | 41 | 378 | 114 | 3 | 3,445 | 124 | 28 | 285 | 74 | 4 | 270 | 76 |
| 2006 | 2,795 | 75 | 37 | 175 | 107 | 2 | 2,757 | 119 | 23 | 251 | 66 | 4 | 252 | 83 |
| 2007 | 1,898 | 80 | 24 | 96 | 75 | 1 | 3,324 | 137 | 24 | 273 | 80 | 3 | 240 | 88 |
| 2008 | 1,133 | 62 | 18 | 54 | 70 | 1 | 1,037 | 126 | 8 | 233 | 67 | 4 | 217 | 71 |
| 2009* | Not | t available |  |  | ot availab |  | 1,280 | 135 | 9 | 293 | 72 | 4 | 191 | 74 |


|  | IR-SSC-7JIrish Scottish Seiners |  |  | IR-SSC-7GIrish Scottish Seiners |  |  | IR-TBB-7JIrish Beam Trawls |  |  | IR-TBB-7GIrish Beam Trawls |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ | Landings | Effort ${ }^{4}$ | LPUE ${ }^{3}$ |
| 1995 | 1,008 | 5 | 192 | 1,123 | 6 | 175 | 0 | 0 | 1 | 63 | 21 | 3 |
| 1996 | 1,100 | 8 | 135 | 1,534 | 10 | 158 | 5 | 1 | 3 | 33 | 27 | 1 |
| 1997 | 806 | 11 | 75 | 2,654 | 16 | 165 | 3 | 2 | 2 | 44 | 28 | 2 |
| 1998 | 467 | 7 | 71 | 2,502 | 15 | 167 | 5 | 5 | 1 | 46 | 35 | 1 |
| 1999 | 77 | 1 | 55 | 1,378 | 8 | 172 | 8 | 7 | 1 | 47 | 41 | 1 |
| 2000 | 187 | 3 | 54 | 1,187 | 10 | 120 | 8 | 7 | 1 | 64 | 37 | 2 |
| 2001 | 236 | 4 | 53 | 1,005 | 16 | 62 | 6 | 3 | 2 | 79 | 40 | 2 |
| 2002 | 409 | 9 | 46 | 1,971 | 21 | 94 | 6 | 3 | 2 | 60 | 32 | 2 |
| 2003 | 371 | 9 | 41 | 1,560 | 21 | 75 | 13 | 9 | 1 | 55 | 49 | 1 |
| 2004 | 314 | 9 | 34 | 1,038 | 19 | 54 | 1 | 2 | 1 | 33 | 55 | 1 |
| 2005 | 253 | 6 | 41 | 1,004 | 15 | 68 | 1 | 2 | 1 | 24 | 50 | 0 |
| 2006 | 192 | 5 | 36 | 912 | 15 | 62 | 1 | 2 | 0 | 19 | 60 | 0 |
| 2007 | 205 | 4 | 58 | 825 | 16 | 52 | 0 | 2 | 0 | 25 | 56 | 0 |
| 2008 | 225 | 3 | 79 | 741 | 12 | 64 | 0 | 1 | 0 | 4 | 37 | 0 |
| 2009* | 347 | 3 | 104 | 731 | 8 | 90 | 0 | 3 | 0 | 2 | 38 | 0 |

1 = LPUE calculated as landings in $\mathrm{kg} / \mathrm{h}$ fishing, power corrected.
$2=$ Effort in hours fishing, power corrected
3 = LPUE calculated as landings in $\mathrm{kg} / \mathrm{h}$ fishing.
4 = Effort in 000 hours fishing.

* Provisional

Table 7.15.10. Whiting in Divisions VIIe-k. XSA Diagnostics.

```
Lowestoft VPA Version 3.1
    6/05/2010 10:28
Extended Survivors Analysis
"Whiting in the Celtic Sea (VIIe-k), WGCSE 2010, COMBSEX (Updated by SD 04/05/20
CPUE data from file whg7ektutrimed.txt
Catch data for 28 years. 1982 to 2009. Ages 0 to 7.
    Fleet, First, Last, First, Last, Alpha, Beta
    year, year, age , age
"FR-GADOID-late: Fre, 1993, 2009, 3, 6, .000, 1.000
"FR-NEPHROPS-Late: F, 1993, 2009, 3, 6, .000, 1.000
"FR-EVHOE: Thalassa , 1997, 2009, 0, 4, .750, 1.000
"UK-WCGFS: UK (E+W), 1987, 2009, 1, 6, .150, . 250
IR-GFS-7G-SweptArea:, 1999, 2009, 0, 6, .750, .920
Time series weights :
    Tapered time weighting not applied
```

Catchability analysis :
Catchability independent of stock size for all ages
Catchability independent of age for ages >= 5
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.000$
Minimum standard error for population
estimates derived from each fleet $=$. 500
Prior weighting not applied
Tuning converged after 28 iterations
Regression weights
, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000
Fishing mortalities

| Age, | 2000, | 2001, | 2002, | 2003, | 2004, | 2005, | 2006, | 2007, | 2008, | 2009 |
| ---: | ---: | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 0, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000 |
| 1, | .037, | .007, | .038, | .016, | .039, | .026, | .053, | .026, | .024, | .004 |
| 2, | .295, | .160, | .168, | .173, | .199, | .285, | .314, | .321, | .131, | .104 |
| 3, | .782, | .686, | .444, | .372, | .481, | .755, | .688, | .922, | .676, | .303 |
| 4, | .967, | 1.272, | .974, | .511, | .597, | 1.006, | .969, | 1.560, | 1.016, | .797 |
| 5, | .972, | 1.387, | 1.050, | .889, | .549, | 1.113, | 1.377, | 1.805, | 1.034, | .717 |
| 6, | .823, | 1.012, | .669, | .828, | .729, | .726, | .993, | 1.858, | 1.201, | .738 |

XSA population numbers (Thousands)

|  |  |  |  | AGE |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | , | 0, | 1, | 2, | 3, | 4, | 5, | 6, |
| 2000 | , | $6.39 E+04$, | 1.10E+05, | 4.05E+04, | 2.10E+04, | 9.75E+03, | 4.18E+03, | 3.93E+03, |
| 2001 | , | 3.88E+04, | 5.23E+04, | 8.69E+04, | 2.47E+04, | 7.88E+03, | 3.04E+03, | 1.29E+03, |
| 2002 | , | 3.84E+04, | 3.18E+04, | 4.25E+04, | 6.06E+04, | 1.02E+04, | 1.81E+03, | 6.21E+02, |
| 2003 |  | 4.27E+04, | 3.14E+04, | 2.50E+04, | 2.94E+04, | 3.19E+04, | 3.14E+03, | 5.18E+02, |
| 2004 |  | 3.98E+04, | 3.50E+04, | 2.53E+04, | 1.72E+04, | 1.66E+04, | 1.56E+04, | 1.06E+03, |


| 2005 |  | 3.54E+04, | 3.26E+04, | 2.75E+04, | 1.70 E+04 | $8.72 \mathrm{E}+03$ | 7.49E+03 | 7.40E+03, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 |  | 5.13E+04, | 2.90E+04, | 2.60E+04, | 1.70E+04, | 6.54E+03, | 2.61E+03, | 2.02E+03, |
| 2007 | , | 7.19E+04, | 4.20E+04, | $2.25 E+04$ | $1.56 \mathrm{E}+04$ | 6.97E+03, | 2.03E+03, | 5.39E+02, |
| 2008 |  | 1.19E+05, | 5.89E+04, | $3.35 E+04$ | $1.34 \mathrm{E}+04$ | 5.06E+03 | 1.20E+03, | 2.74E+02, |
| 2009 |  | 5.60E+04, | 9.74E+04, | 4.71E+04, | 2.40 E | 5.57E+03, | $1.50 \mathrm{E}+03$ | 3.50E+02, |

Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+00,4.59 \mathrm{E}+04,7.94 \mathrm{E}+04,3.47 \mathrm{E}+04,1.45 \mathrm{E}+04,2.05 \mathrm{E}+03,5.99 \mathrm{E}+02$,
Taper weighted geometric mean of the VPA populations:
$6.98 E+04,5.63 E+04,4.24 E+04,2.35 E+04,8.69 E+03,2.46 E+03,5.85 E+02$,
Standard error of the weighted Log(VPA populations) :
.4998, .5147, .5141, .6386, .8797, 1.1387, 1.4433,

Log catchability residuals.
Fleet $: ~ " F R-G A D O I D-l a t e: ~ F r e ~$

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 3, | 4, | 5, | 6 |
| :---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.6427, | -6.1577, | -5.9894, | -5.9894, |
| S.E(Log q), | .5836, | .4297, | .4085, | .3712, |

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 3, | 2.69, | -3.229, | .41, | .21, | 16, | 1.23, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 4, | 1.82, | -4.883, | 3.40, | .72, | 16, | .49, |
| 5, | 1.16, | -1.446, | 5.59, | .85, | 16, | .46, |
| 6, | .95, | .661, | 6.04, | .93, | 16, | .36, |

Fleet : "FR-NEPHROPS-Late: F
Age , 1990, 1991, 1992, 1993, 1994, 1995, 1996, 1997, 1998, 1999 0 , No data for this fleet at this age 1 , No data for this fleet at this age


Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 3, | 4, | 5, | 6 |
| :---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -8.9431, | -8.3682, | -8.1678, | -8.1678, |
| S.E $(\log q)$, | .4791, | .3338, | .3614, | .4467, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time. Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 3, | 1.05, | -.214, | 8.88, | .59, | 16, | .52, | -8.94, |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 4, | 1.01, | -.068, | 8.36, | .84, | 16, | .35, | -8.37, |
| 5, | .89, | 1.414, | 8.20, | .92, | 16, | .31, | -8.17, |
| 6, | .82, | 2.923, | 8.10, | .95, | 16, | .29, | -8.29, |

Fleet : "FR-EVHOE: Thalassa

| Age | , | 1990, | 1991, | 1992, | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | , | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -.50, | -. 21, | 77 |
| 1 | , | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -. 42, | -.83, | 48 |
| 2 | , | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -.61, | -.60, | . 39 |
| 3 | , | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -. 32, | -1.48, | . 29 |
| 4 | , | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -. 08, | -1.22, | 10 |
| 5 |  | No data | for t | is fle | t at | is age |  |  |  |  |  |
| 6 | , | No data | for t | is fle | et at t | is age |  |  |  |  |  |


| Age, | 2000, | 2001, | 2002, | 2003, | 2004, | 2005, | 2006, | 2007, | 2008, | 2009 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0, | -.63, | -.42, | .11, | .20, | -.23, | .33, | -1.15, | 1.02, | 1.23, | -.52 |
| 1, | -.07, | -.12, | -.43, | .90, | 1.43, | -.43, | -.80, | .39, | .01, | -.09 |
| 2, | .51, | .37, | -.40, | 1.04, | 1.19, | -.61, | -.94, | -.56, | -.27, | .47 |
| 3, | -.08, | .73, | -.22, | 1.18, | 1.48, | -.30, | -.82, | -1.08, | .42, | .19 |
| 4, | -.07, | .02, | -.46, | .89, | .86, | .06, | -1.03, | -.40, | .83, | .48 |
| 5 , No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
| 6, No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |

Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4 |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Mean $\log q$, | -6.8422, | -6.9973, | -7.5428, | -7.9747, | -7.9445, |
| S.E(Log q), | .6957, | .6544, | .6938, | .8516, | .6705, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age, Slope , t-value, |  |  |  | Intercept, RSquare, | No Pts, Reg s.e, Mean Q |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0, | .54, | 2.018, | 8.74, | .63, | 13, | .33, | -6.84, |
| 1, | 1.09, | -.181, | 6.65, | .26, | 13, | .75, | -7.00, |
| 2, | .85, | .316, | 7.99, | .28, | 13, | .61, | -7.54, |
| 3, | 1.17, | -.249, | 7.63, | .17, | 13, | 1.03, | -7.97, |
| 4, | 1.05, | -.187, | 7.87, | .54, | 13, | .74, | -7.94, |


| Fleet : "UK-WCGFS: UK (E+W) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | , | 1987, | 1988, | 1989 |  |  |  |  |  |  |  |
| 0 |  | No data | for t | his fle | et at t | this age |  |  |  |  |  |
| 1 | , | -1.22, | -1.40, | -. 18 |  |  |  |  |  |  |  |
| 2 | , | 1.34, | -1.28, | . 05 |  |  |  |  |  |  |  |
| 3 |  | .56, | -.86, | . 40 |  |  |  |  |  |  |  |
| 4 |  | . 06, | -1.11, | . 20 |  |  |  |  |  |  |  |
| 5 |  | 1.13, | .16, | . 62 |  |  |  |  |  |  |  |
| 6 | , | 1.74, | 99.99, | 1.12 |  |  |  |  |  |  |  |
| Age | , | 1990, | 1991, | 1992, | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999 |
| 0 |  | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
| 1 | , | -3.20, | -1.56, | -.14, | 1.29, | 1.62, | -. 52, | -. 10, | . 31, | . 83, | 34 |
| 2 | , | -1.28, | -3.28, | -. 21, | -.16, | . 29, | 1.41, | .69, | .26, | .65, | . 31 |
| 3 | , | -. 21, | -2.38, | -. 12, | . 04, | .55, | .98, | .58, | -.86, | .80, | -. 24 |
| 4 | , | . 08, | -1.37, | .64, | -. 09, | . 44, | .73, | .07, | -. 02, | 51, | -. 49 |
| 5 | , | . 73 , | -.44, | . 39 , | -. 31, | .67, | . 53, | -. 07, | . 08, | -. 75 , | -. 76 |
| 6 | , | .62, | . 24, | . 12 , | .60, | .14, | .88, | -. 58, | -.09, | -1.34, | . 04 |
| Age | , | 2000, | 2001, | 2002, | 2003, | 2004, | 2005, | 2006, | 2007, | 2008, | 2009 |
| 0 |  | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
| 1 |  | 2.49, | 1.44, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |
| 2 |  | .81, | . 39, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |
| 3 |  | . 57, | . 20, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |
| 4 |  | .56, | -. 21, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |
| 5 |  | . 30, | -2.29, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |
| 6 |  | . 41, | -. 18, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age, | 1, | 2, | 3, | 4, | 5, | 6 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -11.3524, | -11.3959, | -11.5834, | -11.6821, | -11.4950, | -11.4950, |
| S.E $(\log q)$, | 1.4620, | 1.1905, | .8605, | .6083, | .8417, | .7964, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 1, | .60, | .856, | 11.29, | .26, | 15, | .88, | -11.35, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2, | .55, | 1.313, | 11.20, | .40, | 15, | .64, | -11.40, |
| 3, | .59, | 2.204, | 11.07, | .69, | 15, | .45, | -11.58, |
| 4, | .91, | .551, | 11.48, | .75, | 15, | .57, | -11.68, |
| 5, | 1.38, | -1.657, | 12.80, | .60, | 15, | 1.09, | -11.50, |

6, 1.44, -2.745, 13.18, .76, 14, .88, -11.23,

Fleet : IR-GFS-7G-SweptArea:

| Age |  | 1990, | 1991, | 1992, | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 33 |
| 1 | 1 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99 | 99.99, | 99.99, | 99.99, | 24 |
| 2 | 2 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | -. 08 |
| 3 | 3 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 11 |
| 4 | 4 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | . 46 |
| 5 | 5 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 29 |
| 6 | 6 | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 99.99, | 09 |


| Age, | 2000, | 2001, | 2002, | 2003, | 2004, | 2005, | 2006, | 2007, | 2008, | 2009 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0, | -.31, | -.07, | -.70, | .10, | .64, | -1.38, | .84, | .07, | .06, | .42 |
| 1, | .26, | -.74, | .04, | -.47, | -.12, | -.14, | .61, | .54, | -.18, | -.03 |
| 2, | .28, | .90, | -.47, | -.45, | -.89, | -.03, | .38, | .75, | .00, | -.40 |
| 3, | -.89, | 1.15, | -.34, | -.79, | -.41, | .10, | .56, | .41, | .39, | -.29 |
| 4, | -1.33, | .98, | -.08, | 99.99, | -.74, | .39, | -.26, | .36, | -.01, | .23 |
| 5 | .01, | .27, | -1.28, | -.77, | -.14, | .89, | 99.99, | 1.27, | -1.29, | .75 |
| 6, | -1.27, | 1.01, | .57, | 99.99, | -2.11, | 1.03, | .80, | .89, | 99.99, | 99.99 |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4, | 5, | 6 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -4.1830, | -4.3005, | -4.7342, | -5.3037, | -5.6291, | -6.3805, | -6.3805, |
| S.E(Log $q)$, | .6261, | .4043, | .5453, | .6132, | .6586, | .8842, | 1.1898, |

Regression statistics :

Ages with q independent of year class strength and constant w.r.t. time.
Age, Slope , t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 0, | .70, | 1.015, | 6.24, | .55, | 11, | .43, | -4.18, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1, | .99, | .026, | 4.35, | .57, | 11, | .42, | -4.30, |
| 2, | .70, | .948, | 6.42, | .53, | 11, | .39, | -4.73, |
| 3, | 1.92, | -1.027, | .99, | .12, | 11, | 1.17, | -5.30, |
| 4, | 1.83, | -.777, | 2.80, | .10, | 10, | 1.23, | -5.63, |
| 5, | .80, | .734, | 6.74, | .64, | 10, | .73, | -6.38, |
| 6, | 1.03, | -.076, | 6.21, | .46, | 8, | 1.32, | -6.25, |

Terminal year survivor and $F$ summaries :
Age 0 Catchability constant w.r.t. time and dependent on age

## Year class $=2009$

| Fleet, | Estimated, Survivors, | Int, | Ext, | Var, Ratio | N, | Scaled, Weights, | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| "FR-GADOID-late: Fre, | 1., | $\begin{aligned} & \text { S.e, } \\ & .000, \end{aligned}$ | $\begin{aligned} & \text { S.e, } \\ & .000, \end{aligned}$ | Ratio, | 0, | Weights, | . 000 |
| "FR-NEPHROPS-Late: F, | 1., | .000, | .000, | .00, | 0, | .000, | 000 |
| "FR-EVHOE: Thalassa, | 27337., | .722, | .000, | .00, | 1, | . 451, | 000 |
| "UK-WCGFS: UK (E+W) | 1., | .000, | .000, | .00, | 0, | .000, | . 000 |
| IR-GFS-7G-SweptArea:, | 70109., | .654, | .000, | .00, | 1, | .549, | . 000 |
| F shrinkage mean | 0., | 1.00, |  |  |  | .000, | . 000 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | R' | Ratio, |  |
| $45857 .$, | .48, | .47, | 2, | .967, | .000 |

Age 1 Catchability constant w.r.t. time and dependent on age
Year class $=2008$

|  | Estimated, | Int, | Ext, | Var, | N, Scaled, | Estimated |  |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet, | Survivors, | s.e, | s.e, | Ratio, | , Weights, | F |  |
| "FR-GADOID-late: Fre, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "FR-NEPHROPS-Late: F, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "FR-EVHOE: Thalassa , | $134933 .$, | .495, | .659, | 1.33, | 2, | .358, | .002 |
| "UK-WCGFS: UK (E+W) , | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| IR-GFS-7G-SWeptArea:, | $79789 .$, | .397, | .044, | .11, | 2, | .555, | .004 |
| F shrinkage mean , | $8942 .$, | $1.00,,,$, |  |  |  | .088, | .034 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | Ratio, |  |  |
| $79443 .$, | .30, | .42, | 5, | 1.434, | .004 |

Age 2 Catchability constant w.r.t. time and dependent on age
Year class $=2007$

| Fleet, | Estimated, Survivors, | Int, | Ext, | Var, Ratio, | $N$ | Scaled, Weights | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| "FR-GADOID-late: Fre, | 1., | .000, | .000, | .00, | 0 | .000, | . 000 |
| "FR-NEPHROPS-Late: F, | 1., | .000, | .000, | .00, | 0 | .000, | 000 |
| "FR-EVHOE: Thalassa | 55965., | .408, | .293, | .72, | 3 | . 363, | . 066 |
| "UK-WCGFS: UK (E+W) | 1., | .000, | .000, | .00, | 0 | . 000, | . 000 |
| IR-GFS-7G-SweptArea:, | 28732., | . 326, | .125, | . 38 , | 3 | . 569, | . 124 |
| F shrinkage mean , | 13285., | 1.00, |  |  |  | . 068, | . 251 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | S.e, | S.e, | ${ }^{\prime}$ | Ratio, |  |
| $34731 .$, | .25, | .20, | 7, | .829, | .104 |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2006$

|  | Estimated, | Int, | Ext, | Var, | N, Scaled, | Estimated |  |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet, | Survivors, | s.e, | s.e, | Ratio, | , Weights, | F |  |
| "FR-GADOID-late: Fre, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "FR-NEPHROPS-Late: F, | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| "FR-EVHOE: Thalassa , | $11760 .$, | .371, | .345, | .93, | 4, | .352, | .363 |
| "UK-WCGFS: UK (E+W) , | $1 .$, | .000, | .000, | .00, | 0, | .000, | .000 |
| IR-GFS-7G-SweptArea:, | $18980 .$, | .291, | .241, | .83, | 4, | .575, | .240 |
| F shrinkage mean , | $4969 .$, | $1.00, \ldots$, |  |  |  | .074, | .711 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | ', | Ratio, |  |
| $14530 .$, | .22, | .22, | 9, | .964, | .303 |

Age 4 Catchability constant w.r.t. time and dependent on age

```
Year class = 2005
```



Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=2004$

| Fleet, | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors, | s.e, | s.e, | Ratio, |  | Weights, | F |
| "FR-GADOID-late: Fre, | 1061., | .413, | .056, | .13, | 2, | .172, | 465 |
| "FR-NEPHROPS-Late: F, | 410., | .385, | .207, | .54, | 2, | .189, | . 930 |
| "FR-EVHOE: Thalassa | 635., | .387, | .393, | 1.02, | 5, | .145, | . 688 |
| "UK-WCGFS: UK (E+W) | 1., | . 000, | .000, | .00, | 0, | . 000, | . 000 |
| IR-GFS-7G-SweptArea:, | 873., | .388, | .158, | .41, | 6, | . 303, | 543 |
| F shrinkage mean | 275., | 1.00, |  |  |  | .191, | 1.190 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Rat | Ratio, |  |
| $599 .$, | .25, | .17, | 16, | .656, | .717 |

Age 6 Catchability constant w.r.t. time and age (fixed at the value for age) 5

```
Year class = 2003
```

| Fleet, | Estimated, Survivors, | Int, | Ext, | Var, Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| "FR-GADOID-late: Fre, | Sur 241. | .400, | .059, | Re .15, | 3, | .258, | . 482 |
| "FR-NEPHROPS-Late: F, | 76., | .390, | .157, | .40, | 3, | .264, | 1.086 |
| "FR-EVHOE: Thalassa | 123., | .364, | .364, | 1.00, | 5, | .052, | . 795 |
| "UK-WCGFS: UK (E+W) | 1. | .000, | .000, | .00, | 0 , | .000, | . 000 |
| IR-GFS-7G-SweptArea:, | 85., | .447, | .341, | . 76, | 6, | .130, | 1.011 |
| F shrinkage mean | 177., | 1.00, |  |  |  | .295, | . 611 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | Ratio, |  |  |
| $137 .$, | .34, | .14, | $18^{\prime}$, | .430, | .738 |

Table 7.15.11. Whiting in Divisions VIIe-k. Fishing mortality (F)-at-age. Fbar range is 2-5.

| AGE | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | $\bigcirc$ | 0 | $\bigcirc$ | 0 |
| 1 | 0.106 | 0.137 | 0.080 | 0.097 | 0.074 | 0.063 | 0.030 | 0.025 | 0.084 | 0.110 |
| 2 | 0.623 | 0.731 | 0.839 | 0.803 | 0.713 | 0.656 | 0.436 | 0.322 | 0.640 | 0.496 |
| 3 | 1.048 | 1.441 | 0.965 | 1.169 | 0.984 | 1.347 | 1.099 | 0.927 | 0.754 | 1.080 |
| 4 | 1.237 | 1.506 | 1.557 | 0.874 | 1.230 | 1.520 | 1.482 | 1.353 | 1.095 | 1.759 |
| 5 | 1.386 | 1.985 | 1.519 | 1.302 | 1.151 | 1.775 | 1.284 | 1.266 | 1.420 | 1.417 |
| 6 | 1.239 | 1.667 | 1.364 | 1.128 | 1.135 | 0.766 | 1.252 | 0.676 | 0.842 | 1.315 |
| +gp | 1.239 | 1.667 | 1.364 | 1.128 | 1.135 | 0.766 | 1.252 | 0.676 | 0.842 | 1.315 |
| FBAR | 1.073 | 1.416 | 1.220 | 1.037 | 1.020 | 1.325 | 1.075 | 0.967 | 0.977 | 1.188 |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 0 | 0 | $\bigcirc$ | 0 | 0 | 0 | $\bigcirc$ | 0 | 0 | $\bigcirc$ | 0 |
| 1 | 0.042 | 0.022 | 0.009 | 0.007 | 0.029 | 0.014 | 0.029 | 0.089 | 0.037 | 0.007 |
| 2 | 0.428 | 0.267 | 0.178 | 0.139 | 0.123 | 0.127 | 0.212 | 0.364 | 0.295 | 0.160 |
| 3 | 0.919 | 0.846 | 0.556 | 0.460 | 0.241 | 0.289 | 0.523 | 0.768 | 0.782 | 0.686 |
| 4 | 0.927 | 0.835 | 0.816 | 0.704 | 0.543 | 0.507 | 0.587 | 1.037 | 0.967 | 1.272 |
| 5 | 1.015 | 1.175 | 0.853 | 0.823 | 0.663 | 0.697 | 0.701 | 1.103 | 0.972 | 1.387 |
| 6 | 0.954 | 1.032 | 1.211 | 0.694 | 0.774 | 0.875 | 0.658 | 0.933 | 0.823 | 1.012 |
| +gp | 0.954 | 1.032 | 1.211 | 0.694 | 0.774 | 0.875 | 0.658 | 0.933 | 0.823 | 1.012 |
| FBAR | 0.822 | 0.781 | 0.601 | 0.531 | 0.393 | 0.405 | 0.506 | 0.818 | 0.754 | 0.876 |
| AGE | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| 0 | 0 | 0 | $\bigcirc$ | 0 | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | 0 |  |  |
| 1 | 0.038 | 0.016 | 0.039 | 0.026 | 0.053 | 0.026 | 0.024 | 0.004 |  |  |
| 2 | 0.168 | 0.173 | 0.199 | 0.285 | 0.314 | 0.321 | 0.131 | 0.104 |  |  |
| 3 | 0.444 | 0.372 | 0.482 | 0.755 | 0.688 | 0.922 | 0.676 | 0.303 |  |  |
| 4 | 0.974 | 0.511 | 0.597 | 1.006 | 0.969 | 1.560 | 1.016 | 0.797 |  |  |
| 5 | 1.050 | 0.889 | 0.549 | 1.113 | 1.377 | 1.805 | 1.034 | 0.718 |  |  |
| 6 | 0.669 | 0.828 | 0.729 | 0.726 | 0.993 | 1.858 | 1.201 | 0.738 |  |  |
| +gp | 0.669 | 0.828 | 0.729 | 0.726 | 0.993 | 1.858 | 1.201 | 0.738 |  |  |
| FBAR | 0.659 | 0.486 | 0.457 | 0.790 | 0.837 | 1.152 | 0.714 | 0.481 |  |  |

Table 7.15.12. Whiting in Divisions VIIe-k. Stock number-at-age ('000).

| AGE | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 62046 | 50256 | 53997 | 71465 | 133033 | 105422 | 33073 | 55008 | 108378 | 163345 |
| 1 | 28885 | 50799 | 41147 | 44209 | 58511 | 108918 | 86313 | 27078 | 45037 | 88732 |
| 2 | 29860 | 21275 | 36282 | 31106 | 32849 | 44494 | 83766 | 68572 | 21625 | 33914 |
| 3 | 16784 | 13116 | 8387 | 12835 | 11408 | 13180 | 18897 | 44351 | 40708 | 9336 |
| 4 | 7108 | 4818 | 2542 | 2617 | 3265 | 3490 | 2806 | 5156 | 14375 | 15677 |
| 5 | 1297 | 1690 | 875 | 439 | 894 | 781 | 625 | 522 | 1092 | 3939 |
| 6 | 64 | 266 | 190 | 157 | 98 | 231 | 108 | 142 | 120 | 216 |
| +gp | 35 | 13 | 41 | 53 | 72 | 122 | 32 | 56 | 40 | 4 |
| TOTAL | 146080 | 142234 | 143460 | 162880 | 240129 | 276639 | 225619 | 200883 | 231374 | 315163 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 0 | 145794 | 193497 | 107334 | 63214 | 58597 | 56795 | 65996 | 134508 | 63904 | 38799 |
| 1 | 133735 | 119366 | 158422 | 87877 | 51755 | 47975 | 46500 | 54033 | 110126 | 52320 |
| 2 | 65102 | 105002 | 95644 | 128554 | 71459 | 41156 | 38728 | 37001 | 40472 | 86928 |
| 3 | 16911 | 34740 | 65811 | 65539 | 91622 | 51744 | 29669 | 25663 | 21044 | 24669 |
| 4 | 2596 | 5523 | 12211 | 30913 | 33885 | 58923 | 31747 | 14394 | 9752 | 7884 |
| 5 | 2210 | 841 | 1962 | 4419 | 12513 | 16115 | 29052 | 14448 | 4178 | 3036 |
| 6 | 782 | 656 | 213 | 684 | 1590 | 5281 | 6569 | 11805 | 3927 | 1294 |
| +gp | 69 | 290 | 273 | 200 | 230 | 1343 | 1997 | 2128 | 3581 | 469 |
| TOTAL | 367200 | 459915 | 441869 | 381402 | 321652 | 279333 | 250257 | 293981 | 256984 | 215400 |
| AGE | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| 0 | 38396 | 42692 | 39835 | 35443 | 51252 | 71912 | 118975 | 56010 | 0 |  |
| 1 | 31766 | 31436 | 34953 | 32614 | 29018 | 41962 | 58877 | 97409 | 45857 |  |
| 2 | 42532 | 25042 | 25320 | 27523 | 26007 | 22522 | 33461 | 47056 | 79443 |  |
| 3 | 60631 | 29428 | 17247 | 16983 | 16952 | 15554 | 13374 | 24039 | 34731 |  |
| 4 | 10175 | 31854 | 16616 | 8725 | 6538 | 6974 | 5063 | 5568 | 14530 |  |
| 5 | 1809 | 3144 | 15647 | 7492 | 2612 | 2031 | 1201 | 1500 | 2054 |  |
| 6 | 621 | 518 | 1059 | 7401 | 2016 | 539 | 274 | 350 | 599 |  |
| +gp | 141 | 39 | 145 | 311 | 1002 | 287 | 48 | 75 | 166 |  |
| TOTAL | 186071 | 164152 | 150820 | 136492 | 135397 | 161781 | 231272 | 232006 | 177380 |  |

Table 7.15.13. Whiting in Divisions VIIe-k. Summary table.

|  | Recruits <br> age 0 |  | TotBIO | TotSSB | Landings | Yield/ssb |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | Fbar 2-5

Table 7.15.14. Whiting in Divisions VIIe-k. Prediction input data.

MFDP version 1a
Run: SFwhg7ekRed2008
Time and date: 14:51 07/05/2010
Fbar age range: 2-5

In puts: F Mean 07-09
Catch and stock weights are mean 07-09
Recruits age 0 in 10, 11 and 12 GM mean 82-09 (2008 reduced by 25\%)

| 2010 |  |  |  | PF |  | PM | SWt |  | Sel | CWt |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 69136 | 0.2 | 0.00 |  | 0 | 0 | 0.000 |  | 0.000 | 0.039 |
|  | 1 | 45857 | 0.2 | 0.39 |  | 0 | 0 | 0.217 |  | 0.018 | 0.229 |
|  | 2 | 59582 | 0.2 | 0.90 |  | 0 | 0 | 0.290 |  | 0.185 | 0.301 |
|  | 3 | 34731 | 0.2 | 0.99 |  | 0 | 0 | 0.393 |  | 0.634 | 0.391 |
|  | 4 | 14530 | 0.2 | 0.99 |  | 0 | 0 | 0.547 |  | 1.124 | 0.503 |
|  | 5 | 2054 | 0.2 | 1.00 |  | 0 | 0 | 0.652 |  | 1.185 | 0.604 |
|  | 6 | 599 | 0.2 | 1.00 |  | 0 | 0 | 0.721 |  | 1.265 | 0.662 |
|  | 7 | 166 | 0.2 | 1.00 |  | 0 | 0 | 0.951 |  | 1.265 | 0.900 |




Input units are thousands and kg - output in tonnes

Table 7.15.15. Whiting in Divisions VIIe-k. Management options table.

## MFDP version 1a

Run: SFwhg7ekRed2008 (2008 yearclass strenght reduced by 25\%)
Whiting in the Celtic Sea (VIIe-k), WGCSE 2009, COMBSEX (Updated by SD 04/05/2010)
Time and date: 14:51 07/05/2010
Fbar age range: 2-5
$\left.\begin{array}{ccccccc}\begin{array}{c}\text { 2010 } \\ \text { Biomass }\end{array} & \text { SSB } & \text { FMult } & \text { FBar } & \text { Landings }\end{array}\right)$

[^17]Table 7.15.16. Whiting in Divisions VIIe-k. Detailed results.

MFDP version 1a
Run: SFwhg7ekRed2008 (2008 yearclass strenght reduced by 25\%)
Time and date: 14:51 07/05/2010
Fbar age range: 2-5


Input units are thousands and kg - output in tonnes

Table 7.15.17. Whiting in Divisions VIIe-k. Stock numbers of recruits and the source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes.


Whiting VIle-k : Year-class \% contribution to


Table 7.15.18. Whiting VIIe--k (Celtic Sea) input ".sen" data file for maximum sustainable yield analysis.


Table 7.15.19. Whiting VIIe-k (Celtic Sea) input ".sum" data file for maximum sustainable yield analysis.

```
Stock summary, Whiting ,VIIe-k
            12
    Year
        1982 2009
Recruits, age 0, (millions)
        0 1000000
SSB, ('000 t)
        1000
TSB, ('000 t)
        1000
Catch, Total ('000 t)
        1000
Catch, H.cons ('000 t)
        1000
Not used
        1000
Not used
        1000
Mean F, Total
            2 5
Mean F, H.cons.
Not used
            0 0
Not used
\begin{tabular}{lrrlllllllll}
1982 & & 62 & 19.0 & 22.6 & 11.2 & 11.2 & 0.0 & 0.0 & 1.073 & 1.073 & 0.000 \\
1983 & 50 & 17.0 & 22.8 & 11.8 & 11.8 & 0.0 & 0.000 & 1.416 & 1.416 & 0.000 & 0.000 \\
1984 & 54 & 17.5 & 23.4 & 10.0 & 10.0 & 0.0 & 0.0 & 1.220 & 1.220 & 0.000 & 0.000 \\
1985 & 71 & 17.6 & 23.3 & 10.8 & 10.8 & 0.0 & 0.0 & 1.037 & 1.037 & 0.000 & 0.000 \\
1986 & 133 & 18.6 & 26.1 & 10.0 & 10.0 & 0.0 & 0.0 & 1.020 & 1.020 & 0.000 & 0.000 \\
1987 & 105 & 25.0 & 37.6 & 12.7 & 12.7 & 0.0 & 0.0 & 1.325 & 1.325 & 0.000 & 0.000 \\
1988 & 33 & 33.8 & 45.8 & 15.1 & 15.1 & 0.0 & 0.0 & 1.075 & 1.075 & 0.000 & 0.000 \\
1989 & 55 & 34.8 & 39.5 & 16.5 & 16.5 & 0.0 & 0.0 & 0.967 & 0.967 & 0.000 & 0.000 \\
1990 & 108 & 27.5 & 32.8 & 14.1 & 14.1 & 0.0 & 0.0 & 0.977 & 0.977 & 0.000 & 0.000 \\
1991 & 163 & 24.2 & 33.3 & 13.5 & 13.5 & 0.0 & 0.0 & 1.188 & 1.188 & 0.000 & 0.000 \\
1992 & 146 & 32.3 & 48.2 & 12.4 & 12.4 & 0.0 & 0.0 & 0.822 & 0.822 & 0.000 & 0.000 \\
1993 & 193 & 47.1 & 61.9 & 16.3 & 16.3 & 0.0 & 0.0 & 0.781 & 0.781 & 0.000 & 0.000 \\
1994 & 107 & 62.6 & 82.2 & 20.0 & 20.0 & 0.0 & 0.0 & 0.601 & 0.601 & 0.000 & 0.000 \\
1995 & 63 & 74.6 & 84.1 & 22.7 & 22.7 & 0.0 & 0.0 & 0.531 & 0.531 & 0.000 & 0.000 \\
1996 & 59 & 72.7 & 79.3 & 18.3 & 18.3 & 0.0 & 0.0 & 0.393 & 0.393 & 0.000 & 0.000 \\
1997 & 57 & 62.9 & 67.6 & 20.5 & 20.5 & 0.0 & 0.0 & 0.405 & 0.405 & 0.000 & 0.000 \\
1998 & 66 & 50.0 & 55.3 & 19.2 & 19.2 & 0.0 & 0.0 & 0.506 & 0.506 & 0.000 & 0.000 \\
1999 & 135 & 39.4 & 44.1 & 19.9 & 19.9 & 0.0 & 0.0 & 0.818 & 0.818 & 0.000 & 0.000 \\
2000 & 64 & 34.6 & 45.4 & 14.9 & 14.9 & 0.0 & 0.0 & 0.754 & 0.754 & 0.000 & 0.000 \\
2001 & 39 & 39.9 & 48.1 & 12.8 & 12.8 & 0.0 & 0.0 & 0.876 & 0.876 & 0.000 & 0.000 \\
2002 & 38 & 40.2 & 45.2 & 13.1 & 13.1 & 0.0 & 0.0 & 0.659 & 0.659 & 0.000 & 0.000 \\
2003 & 43 & 33.5 & 37.9 & 10.6 & 10.6 & 0.0 & 0.0 & 0.486 & 0.486 & 0.000 & 0.000 \\
2004 & 40 & 30.4 & 34.7 & 10.0 & 10.0 & 0.0 & 0.0 & 0.457 & 0.457 & 0.000 & 0.000 \\
2005 & 35 & 27.0 & 31.0 & 12.0 & 12.0 & 0.0 & 0.0 & 0.789 & 0.789 & 0.000 & 0.000 \\
2006 & 51 & 22.4 & 26.4 & 9.5 & 9.5 & 0.0 & 0.0 & 0.837 & 0.837 & 0.000 & 0.000 \\
2007 & 72 & 21.2 & 27.2 & 8.9 & 8.9 & 0.0 & 0.0 & 1.152 & 1.152 & 0.000 & 0.000 \\
2008 & 119 & 23.2 & 32.5 & 5.7 & 5.7 & 0.0 & 0.0 & 0.714 & 0.714 & 0.000 & 0.000 \\
2009 & 56 & 33.7 & 48.2 & 5.7 & 5.7 & 0.0 & 0.0 & 0.480 & 0.480 & 0.000 & 0.000
\end{tabular}
```

Table 7.15.20. Whiting VIIe-k (Celtic Sea) output table from maximum sustainable yield analysis.

## Ricker

329/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB <br> Alpha | ADMB <br> Beta | Unscaled <br> Alpha | Unscaled <br> Beta | AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Deterministic | 5.00 | 0.90 | 38.10 | 14.99 | 1.49 | 0.50 | 5.55 | 0.03 | 48.02 |
| Mean | 2.37 | 0.55 | 41.54 | 15.53 | 1.49 | 0.50 | 5.63 | 0.03 |  |
| 5\%ile | 0.89 | 0.29 | 27.95 | 10.92 | 1.16 | 0.27 | 3.50 | 0.01 |  |
| 25\%ile | 1.52 | 0.41 | 33.03 | 13.24 | 1.34 | 0.41 | 4.62 | 0.02 |  |
| 50\%ile | 2.13 | 0.52 | 38.66 | 15.23 | 1.47 | 0.51 | 5.44 | 0.03 |  |
| 75\%ile | 3.18 | 0.66 | 45.02 | 17.07 | 1.61 | 0.58 | 6.42 | 0.03 |  |
| 95\%ile | 4.46 | 0.94 | 66.33 | 21.79 | 1.85 | 0.70 | 8.36 | 0.04 |  |
| CV | 0.47 | 0.37 | 0.34 | 0.22 | 0.15 | 0.26 | 0.26 | 0.26 |  |
|  |  |  |  |  |  |  |  |  |  |

Beverton-Holt
0/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB <br> Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Deterministic | 5.00 | 0.81 | 36.53 | 13.79 | 0.61 | 0.61 | 70.71 | 0.00 | 46.65 |
| Mean |  |  |  |  |  |  |  |  |  |
| 5\%ile |  |  |  |  |  |  |  |  |  |
| 25\%ile |  |  |  |  |  |  |  |  |  |
| 50\%ile |  |  |  |  |  |  |  |  |  |
| 75\%ile |  |  |  |  |  |  |  |  |  |
| 95\%ile |  |  |  |  |  |  |  |  |  |
| CV |  |  |  |  |  |  |  |  |  |

## Smooth hockeystick

559/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB <br> Alpha | ADMB <br> Beta | Unscaled <br> Alpha | Unscaled <br> Beta | AIC |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |

Table 7.15.21. Whiting in Divisions VIIe-k. Yield-per-recruit summary table.

MFYPR version 2a
Run: YPRwhg7ekRed2008 (2008 yearclass strenght reduced by 25\%)
Time and date: 14:52 07/05/2010

| Yield per results <br> FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 2.5837 | 3.9402 | 2.4513 | 3.9402 | 2.4513 |
| 0.1000 | 0.0782 | 0.1983 | 0.1250 | 4.5304 | 1.7059 | 2.9546 | 1.5738 | 2.9546 | 1.5738 |
| 0.2000 | 0.1564 | 0.2887 | 0.1662 | 4.0826 | 1.3302 | 2.5074 | 1.1983 | 2.5074 | 1.1983 |
| 0.3000 | 0.2347 | 0.3417 | 0.1831 | 3.8218 | 1.1244 | 2.2471 | 0.9928 | 2.2471 | 0.9928 |
| 0.4000 | 0.3129 | 0.3772 | 0.1907 | 3.6483 | 0.9955 | 2.0741 | 0.8641 | 2.0741 | 0.8641 |
| 0.5000 | 0.3911 | 0.4029 | 0.1944 | 3.5228 | 0.9073 | 1.9490 | 0.7762 | 1.9490 | 0.7762 |
| 0.6000 | 0.4693 | 0.4228 | 0.1963 | 3.4267 | 0.8432 | 1.8534 | 0.7122 | 1.8534 | 0.7122 |
| 0.7000 | 0.5476 | 0.4387 | 0.1972 | 3.3500 | 0.7942 | 1.7772 | 0.6634 | 1.7772 | 0.6634 |
| 0.8000 | 0.6258 | 0.4519 | 0.1976 | 3.2869 | 0.7554 | 1.7145 | 0.6248 | 1.7145 | 0.6248 |
| 0.9000 | 0.7040 | 0.4630 | 0.1977 | 3.2337 | 0.7239 | 1.6616 | 0.5935 | 1.6616 | 0.5935 |
| 1.0000 | 0.7822 | 0.4727 | 0.1977 | 3.1879 | 0.6976 | 1.6162 | 0.5673 | 1.6162 | 0.5673 |
| 1.1000 | 0.8605 | 0.4811 | 0.1976 | 3.1479 | 0.6752 | 1.5766 | 0.5451 | 1.5766 | 0.5451 |
| 1.2000 | 0.9387 | 0.4886 | 0.1974 | 3.1125 | 0.6559 | 1.5415 | 0.5259 | 1.5415 | 0.5259 |
| 1.3000 | 1.0169 | 0.4954 | 0.1972 | 3.0808 | 0.6390 | 1.5101 | 0.5092 | 1.5101 | 0.5092 |
| 1.4000 | 1.0951 | 0.5015 | 0.1970 | 3.0521 | 0.6240 | 1.4818 | 0.4943 | 1.4818 | 0.4943 |
| 1.5000 | 1.1734 | 0.5070 | 0.1967 | 3.0260 | 0.6106 | 1.4560 | 0.4811 | 1.4560 | 0.4811 |
| 1.6000 | 1.2516 | 0.5121 | 0.1965 | 3.0021 | 0.5986 | 1.4324 | 0.4691 | 1.4324 | 0.4691 |
| 1.7000 | 1.3298 | 0.5169 | 0.1962 | 2.9800 | 0.5876 | 1.4106 | 0.4583 | 1.4106 | 0.4583 |
| 1.8000 | 1.4080 | 0.5213 | 0.1960 | 2.9596 | 0.5776 | 1.3905 | 0.4484 | 1.3905 | 0.4484 |
| 1.9000 | 1.4863 | 0.5254 | 0.1957 | 2.9405 | 0.5684 | 1.3717 | 0.4393 | 1.3717 | 0.4393 |
| 2.0000 | 1.5645 | 0.5292 | 0.1955 | 2.9227 | 0.5599 | 1.3541 | 0.4309 | 1.3541 | 0.4309 |


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(2-5) | 1.0000 | 0.7822 |
| FMax | 0.9342 | 0.7308 |
| F0.1 | 0.2111 | 0.1651 |
| F35\%SPR | 0.406 | 0.3176 |

Weights in kilograms

Irish landings for the main gear types by quarter in 2009:


UK (E\&W) whiting landings for all gears 2009:


Figure 7.15.1. Whiting in VIIe-k (Celtic Sea). The spatial and temporal distribution of landings data in 2009 available to the WG. French landings distributions for otter trawlers and twin rigged otter trawlers was not available to the WG in 2010.


Figure 7.15.2. Whiting in VIIe-k (Celtic Sea). 2009 length compositions (raised numbers) of French, UK and Irish fleets.
(a)

(b)


Figure 7.15.3. Whiting in VIIe-k (Celtic Sea). Annual landings age composition (a) and standardized catch proportions-at-age (b).


Figure 7.15.4. Whiting in VIIe-k (Celtic Sea). Stock weights-at-age.

VIIe-k UK, 96 trips, 1088 hauls



VIIj Ireland Otter trawl, 15 trips, 200 hauls


Figure 7.15.5. Whiting in VIIe-k (Celtic Sea). 2009 Annual length compositions of Irish, UK and French discards. Numbers are raised to the sampled catch for the UK and are raised by trip to the fleet for Ireland and are unraised sampled lengths for France.


Figure 7.15.6. Whiting in VIIe-k (Celtic Sea). Age Composition of Discards from Irish Otter board trawlers 2003-2009 in VIIg (left) and VIIj (right).
(a)


FRGADOIDlate

log index
(b)

log index
frevhoe

(b) Cont.

## IRGFS7GSweptArea



Figure 7.15.7. Whiting in VIIe-k (Celtic Sea). Pair wise scatterplots for the log numbers-at-age for the main tuning fleets to examine internal constancy of the indices (a) commercial fleets and (b) surveys.
(a)

(b)


Figure 7.15.8. Whiting in VIIe-k (Celtic Sea). Mean log standardized plots of indices by (a) age and year, and (b) age and cohort.


Figure 7.15.9. Whiting in VIIe-k (Celtic Sea). (a) standardized and (b) $\log$ standardized plots of survey indices used within the assessment for younger ages (0-3) by cohort.


Figure 7.15.10. Whiting in VIIe-k (Celtic Sea). Landings, Effort and Landings per Unit of Effort (lpue) for some fleets landing whiting. For the UK fleets Effort is GRT corrected.

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Figure 7.15.11. Whiting in VIIe-k (Celtic Sea). The survivor estimate weightings given by all fleets.
(a)

(b)


Figure 7.15.12. Whiting in VIIe-k (Celtic Sea). Log fleet catchability residuals bubble (a) and line (b) plots.


Figure 7.15.13. Whiting in VIIe-k (Celtic Sea). Retrospective analysis.


Figure 7.15.14. Whiting in VIIe-k (Celtic Sea). Stock summary.


Figure 7.15.16. Whiting in VIIe-k (Celtic Sea). Stock-recruitment relationship.


Figure 7.15.17. Whiting in VIIe-k (Celtic Sea). Fitted stock recruit-relationships with 1000 MCMC resamples. Left hand panels illustrate confidence intervals for converged estimates of $\mathbf{F}_{\text {msy }}$. Right hand panels present curves plotted from the first 100 converged MCMC re-samples for illustration. The blue line indicates the deterministic estimate. The legends for each recruitment model show the number of converged values of $\mathrm{F}_{\text {msy }}$ from the $\mathbf{1 0 0 0}$ re-samples.
(a)

Whg-7e-k Ricker

(b)


Figure 7.15.18. Whiting in VIIe-k (Celtic Sea). Estimates of $\mathbf{F}$ reference points and equilibrium yield and SSB plots. Left hand panels illustrate confidence intervals for converged estimates. Right hand panels present curves plotted from the first 100 converged MCMC re-samples for illustration. The blue line indicates the deterministic estimate. Circles show assessment estimates with the most recent year labelled. (a) Ricker and (b) Hockey stick stock-recruitment models.


Figure 7.15.19. Whiting in VIIe-k (Celtic Sea). Fitted F reference points, yield-per-recruit and SSB per recruit against mortality with confidence intervals estimated by parametric re-sampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v.


MFYPR version 2a
Run: YPRwhg7ekRed2008
Time and date: 14:52 07/05/2010

| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(2-5) | 1.0000 | 0.7822 |
| FMax | 0.9342 | 0.7308 |
| F0.1 | 0.2111 | 0.1651 |
| F35\%SPR | 0.4060 | 0.3176 |

MFDP version 1a
Run: SFwhg7ekRed2008
Whiting in the Celtic Sea (VIIe-k), WGCSE 2009, COMBSEX (Updated by SD 04/05/2010) Time and date: 14:51 07/05/2010
Fbar age range: $2-5$
Input units are thousands and kg - output in tonnes

Weights in kilograms

Figure 7.15.20. Whiting in VIIe-k (Celtic Sea). Yield-per-recruit and short-term forecast plots.

### 7.16 Whiting in Divisions VIIb, c

## Type of assessment

No assessment.
The nominal landings are given in Table 7.16.1.

Table 7.16.1. Nominal Landings ( $\mathbf{t}$ ) of Whiting in Division VIIb,c for 1995-2009.

| Country | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ |  |  |  |  |  |  |  |  |  |  |  |  |
| France | 57 | 76 | 65 | $37^{*}$ | $\ldots 1^{*}$ | 107 | 114 | 111 | 92 | 59 | 102 | 62 | 32 |
| Ireland | 1,894 | 1,233 | 403 | 323 | 206 | 563 | 357 | 386 | 423 | 135 | 65 | 49 | 100 |
| a | 76.0 | 76.3 |  |  |  |  |  |  |  |  |  |  |  |
| Netherlands | - | - | - | - | - | - | 2 | - | 3 | - | 2 | - | - |
| Spain | + | + | - | 27 | 1 | 4 | - | 6 | - | 31 | 18 | 19 | 1 |
| 4 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK(E/W/NI) | 24 | 96 | 75 | 49 | 10 | 6 | 5 | 4 | 5 | 1 | 11 | 5 | 1 |
| UK(Scotland) | 71 | 17 | 4 | 27 | - | 19 | 1 | + | - | - | - | - | - |
| Total | 2,046 | 1,422 | 547 | 463 | 217 | 699 | 479 | 507 | 523 | 226 | 198 | 135 | 134 |

*See VIIg-k.
${ }^{a}$ provisional

### 8.1 Western Channel overview

There is no overview.

### 8.2 Plaice in the Western Channel (ICES Divisions VIIe)

Type of assessment in 2010
Update assessment with no changes to the assessment settings agreed at the Benchmark Assessment meeting (WKFLAT 2010) in February 2010.

WKFLAT had considerable debate over the most appropriate assessment model. WKFLAT felt that the biologically based approach, which incorporates the migration issue by including $15 \%$ of VIId quarter 1 catches in the assessment for VIIe is a sound basis for advice. However it recognised that it is dependent on the assumption that historical patterns of migration have persisted and that the relative size of the subpopulations has been roughly stable. WKFLAT recognised that there were possible implementation issues which could not be reviewed by the meeting with regards to keeping other assessments, such as North Sea plaice, consistent with the information used. In addition, WKFLAT was unable to anticipate all of the management problems that may arise from such a change. Therefore a 'back-up' alternative assessment methodology based on VIIe catches only was developed (the 'truncated model').

Given the improved performance of the assessment with regards to the retrospective pattern, WKFLAT recommended that the historical short-term forecast methodology be reinstated; this was previously removed because of retrospective bias in F and SSB in the assessment.

This recommended model differed from that previously used at the Working Group:

- $15 \%$ of quarter 1 catches (and associated age compositions) from UK (E\&W), Belgium and France added into VIIe data;
- First year of catch and weight-at-age data changed from 1976 to 1980;
- UK (E\&W) FSP survey data truncated to exclude age 9;
- $F($ Bar $)$ age-range reduced from $F(3-7)$ to $F(3-6)$.

ICES advice applicable to 2009
Exploitation boundaries in relation to precautionary limits: Given the low stock size, recent poor recruitment, high fishing mortality, the uncertainty in the assessment, and the inability to reliably forecast catch, ICES recommends a substantial reduction in catch until the estimate of SSB is above $B_{p a}$ or other strong evidence of rebuilding is observed.

## ICES advice applicable to 2010

Exploitation boundaries in relation to precautionary limits: Given the low stock size, recent poor recruitment, high fishing mortality, the uncertainty in the assessment, and the inability to reliably forecast catch, ICES recommends a substantial reduction in catch until the estimate of SSB is above $B_{p a}$ or other strong evidence of rebuilding is observed.

### 8.2.1 General

## Stock description and management units

The management area for this stock is strictly that for ICES Area VIIe called the Western English Channel. The TAC area does not correspond to the stock area as it includes the larger component of VIId (Eastern English Channel). However as determined by WKFLAT 2010, a significant proportion of the catches of the VIIe stock are taken in the adjacent area during the time of spawning. Plaice is not the target species in VIIe, and it is generally caught as a bycatch by the sole and anglerfish directed fleets.


## Management applicable to 2009 and 2010

There are technical measures in operation including a minimum 80 mm mesh size and a MLS $(27 \mathrm{~cm})$ for this species.

The TAC and the national quotas by country for 2009

| Species: Plaice <br> Pleuronectes platess $a$ | Zone:VIId and VIIe <br> (PLE/7DE.) |
| :--- | :--- | :--- |
| Belgium 760 <br> France 2534 <br> United Kingdom 1352 <br> EC 4646 <br> TAC 4646 | Analytical TAC <br> Article 3 of Regulation (EC) No 847/96 <br> applies. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

In addition, Annex IIc, restricts the number of days-at-sea to 192 for beam trawlers of mesh size equal to or greater than 80 mm , and for static nets including gillnets, trammelnets and tanglenets, with mesh size less than 220 mm , with an additional 12 days for the UK beam trawl fleet due to a reduction in capacity of the fleet.

The TAC and the national quotas by country for 2010

| Species:Plaice <br> Pleuronectes platessa | Zone:VIId and VIle <br> (PLE/7DE.) |  |
| :--- | :--- | :--- | :--- |
| Belgium | 699 |  |
| France | 2332 |  |
| United Kingdom | 1243 |  |
| EU | 4274 | Analytical TAC |
| TAC | 4274 |  |

In addition, Annex IIc, restricts the number of days-at-sea to 164 for beam trawlers of mesh size equal to or greater than 80 mm , and for static nets including gillnets, trammelnets and tanglenets, with mesh size less than 220 mm , with an additional 12 days for the UK beam trawl fleet due to a reduction in capacity of the fleet.

## The fishery in 2009

A full description of the fishery is provided in the Stock Annex, Section A2.
In the western English Channel plaice are taken mainly as a bycatch in beam trawls directed at sole and anglerfish. In 2009, the UK beam trawl fleet took around $57 \%$ of the total landing of this stock with the UK otter trawl fleet taking around $21 \%$. The remainder of the landings is taken by the French and Belgian fleets.

UK Otter trawl effort in 2009 continues the downward trend whereas the UK beam trawl effort has now fallen sharply from the high level observed over the period 2003-2008, and is now at the same level observed in 2000.

This stock is the smaller of the two stocks that make up the larger TAC area of VIId,e. The landings from this stock in 2009 and 2008 amounted to around $20 \%$ of the TAC.

## Landings

National landings data reported to ICES, and estimates of total landings used by the Working Group, are given in Table 8.2.1. Estimated total international landings in 2009 were 916 t . The Working Group estimate of the 2008 landings was revised upwards due to minor revisions to the landings by UK (E\&W) and UK (Guernsey) but these had minimal impact.

Landings increased to levels of 2600 t during the latter half of the 1980 s due to a series of good recruitments in 1986-1988, but subsequently dropped to levels fluctuating around 1200 t . The last three years have seen landings fall to under 1000 t . Unallocated landings in recent years, are generally the additional French landings derived from sales note information. In addition to the reported landings for VIIe, an extra 127 tonnes was added from the VIId plaice stock representing an adjustment for migration of $15 \%$ of quarter 1 between the two stocks. This process was agreed at the Benchmark Assessment meeting in February 2010 and the method is documented in the Stock Annex. A reciprocal correction was made to the VIId stock.

## Data

Sampling levels are detailed Section 2 (Table 2.1).

## Discards

Discards estimates, from the UK (E\&W) and French discard sampling programme, are available for the period 2002-2009 (Annual Data Files on ICES network) and indicate that discarding appears to be higher in quarters 1 and 2 in this fishery, but is still low compared to other plaice stocks. Quarterly profiles of numbers landed and dis-carded-at-length, in 2009, are given in Figure 8.2.2.

## Biological

Annual length compositions of the UK (E\&W) landings in 2009 are provided for two UK fleets (Table 8.2.3). No length data for the French landings were available. Length distributions of UK (E\&W) landings from 2000 to 2009 as used by the WG are illustrated in Figure 8.2.3.

Quarterly age compositions for landings in 2009 were available from UK (E\&W) only, which accounted for almost $80 \%$ of the total reported international landings. Additional age compositions representing the migration adjustment ( $15 \%$ of quarter 1 landings for VIId) were available from UK (E\&W), Belgium and France. The method for the derivation of the international catch numbers and the calculation of the catch and stock weights-at-age are fully described in the Stock Annex, Section B1. Catch numbers-at-age landed annually (including migration element) are given in Table 8.2.4 and plotted for 2000 to 2009 in Figure 8.2.4. Catch and stock weights-at-age are given in Tables 8.2.5 and 8.2.6.

Catch weights are plotted as mid-year values; stock weights are interpolated back to January 1st, as standard for this stock. The standard settings used for natural mortality and the proportions of F and M before spawning were used. (See Stock Annex).

## Surveys

There are currently two surveys that provide abundance estimates to the Working Group. The UK (E\&W) commercial beam trawl survey has used the FV Carhelmar for most survey years with the exception being 2002 and 2004, when the RV CORYSTES was used instead. Detailed information on the survey protocols and area coverage can be found in the Stock Annex. Table 8.2.7 gives abundance indices as numbers caught per 100 km for age groups 1 to 9 as obtained by UK-WECBTS. Strong and weak year classes have been well tracked by this survey in the past. (Figure 8.2.6).

Since 2003 the UK Fisheries Science Partnership (FSP: Cefas-UK industry cooperative project) has been conducting a survey using commercial vessels with scientific observers and following a standard grid of stations extending from the Scilly Isles to Lyme Bay. The survey covers a substantially larger area than the current survey (UKWECBTS) and is thought to be more representative of the stock in UK waters. This dataset was first included in the 2007 assessment, and the exploratory analysis can be seen in that report (ICES, 2007; Section 3.2.5). There have been a number of vessel changes, gear changes and temporal variations in this survey series, but the survey has performed well in tracking year classes in the past. However, a strong year effect was noticed in 2008 having a significant impact on the survivor estimates. The 2009 WG excluded the 2008 data.

## Commercial fleet effort and cpue

The UK (E\&W) cpue data shows the individual fleets that make up the composite of all otter trawl and all beam trawl fleets that are used in the commercial tuning data sets. Trends in lpue and effort are given in Table 8.2.2 and Figure 8.2.1; more detailed
information on the distribution of effort by area and trends in the fishery can be found in the Stock Annex. Lpue in the North of VIIe for both commercial beam and otter trawlers reached a peak in 1988-1990, fell sharply to 1995 and is now at stable but low levels. Survey cpue (Beam trawl survey in the North of VIIe) has shown a similar but slightly earlier trend in the early years but indicates a more pronounced temporary increase in catches during 2000 and 2001 compared to the commercial series. Commercial beam trawl lpue in the South and West of VIIe shows a general decline from 1990 to 2008, with otter trawler lpue declining slowly since 1997 in the west, but showing much more variation throughout the time-series in the south. All lpue time-series show an increase in 2009.

Effort (fishing power corrected, using GRT) by UK (E\&W) beam trawlers shows an increasing trend between 1992 and 2003, then remaining stable at this high level until 2008. In 2009 effort fell dramatically back to the level observed in 2000. In contrast, effort by otter trawlers continues to decline slowly from the highest values shown at the beginning of the time-series.

### 8.2.2 Stock assessment

## Catch-at-age analysis

Section 1.3 outlines the general approach adopted at this year's Working Group meeting, and the specific approach for this stock is given in the Stock Annex. All relevant tuning and XSA outputs not included in this report are available in the 'Exploratory runs' folder. The details of the previous assessment approaches for this stock can be found in the Stock Annex.

## Data screening

The age range for the analysis was $1-10+$, as standard.
As this was an update assessment, full data screening, tuning data and exploratory XSA trials were not carried out. For catch data screening, a separable VPA was carried out using the standard setting as detailed in the Stock Annex. The results (Figure 8.2.5.cont.) show no anomalies in recent years, and high residuals on the youngest age as previously observed. The changes made to the assessment data as recommended at WKFLAT made no improvement.

Tuning information available consisted of same five fleets as last year: three UK commercial series, UK otter historic, UK otter trawl, UK beam trawl; and two UK survey-series: UK-WEC-BTS, and UK (E+W) FSP. These are presented in Table 8.2.8. The figures in bold indicate the data used for the final run.

Details of the derivation of the tuning fleets are presented in the Stock Annex.
Tuning indices were examined for inconsistencies using SURBA version 3.0. $\log$ (cpue) plots plotted by year class and by year (Figure 8.2.6). Four of the tuning indices indicate highly consistent year-class estimates, and plots of index by year do not indicate substantial year effects in the tuning data. The UK (E\&W) FSP indices show a large year effect in the 2008 data. Inclusion of these data at the WGCSE 2009 led to the final estimates of each year-class for this fleet being reduced significantly from the previous year's estimate at all ages and given that this fleet's estimates received heavy weighting in the final estimates or survivors, this data was excluded from the final assessment. There were a number of changes to the survey in 2008, but these mostly affected the eastern part of the survey, whereas the greatest change in abundance was noted in the western survey and these changes continued in 2009.

The addition of the 2009 data for this survey did not eradicate this problem, but greatly reduced its impact. Nevertheless the 2008 data have been excluded once again. The cause of this year effect remains unclear.

In addition, this dataset requires further analysis and standardisation across years. Also consideration should be given to using the standard 12 m beams on future surveys in order to ensure consistency in the gear selectivity.

## Final update assessment

The settings used for the final run are shown in the table. The full assessment history is given in the Stock Annex.

|  |  | 2009 XSA | 2010 WKFLAT | 2010 XSA |
| :---: | :---: | :---: | :---: | :---: |
| Catch-at-age data |  | $\begin{gathered} \text { 1976-2008, 1- } \\ 10+ \end{gathered}$ | 1980-2008, 1-10+ <br> add catch from 7d | 1980-2009, 1-10+ <br> add catch from 7d |
| Fleets | UK-WECBTS - Survey | 1986-08, 1-8 | 1986-08, 1-8 | 1986-09, 1-8 |
|  | UK WECOT - Commercial | 1988--08, 3-9 | 1988-08, 3-9 | 1988-09, 3-9 |
|  | UK WECOT-Commercial historic | 1976-87, 2-9 | 1980-87, 2-9 | 1980-87, 2-9 |
|  | UK WECBT - Commercial | 1989-08, 3-9 | 1989-08, 3-9 | 1989-09, 3-9 |
|  | UK E+W FSP - Survey | 2003-07, 2-9 | 2003-07, 2-8 | $\begin{gathered} \text { 2003-09, 2-8 (exc } \\ 08 \text { ) } \end{gathered}$ |
| Taper |  | No | No | No |
| Taper range |  | - | - | - |
| Ages catch dep. Stock size |  | None | None | None |
| q plateau |  | 7 | 7 | 7 |
| F shrinkage se |  | 2.5 | 2.5 | 2.5 |
| year range |  | 5 | 5 | 5 |
| age range |  | 4 | 4 | 4 |
| Fleet SE threshold |  | 0.5 | 0.5 | 0.5 |
| Prior weighting |  | - | - | - |
| Plus group |  | 10 | 10 | 10 |
| F Bar Range |  | F(3-7) | F(3-6) | F(3-6) |

The diagnostics for the final XSA run are shown in Table 8.2.9 and the catchability residuals are plotted in Figure 8.2.5. Some weak trends/patterns can be seen in the commercial beam trawl and otter trawl fleets (UK-WECBT; UK-WECOT) and a year effect can be seen in the survey results (UK-WEC-BTS) for 2004 probably associated with a change in vessel effect.

Estimates for the youngest ages are almost entirely determined by the UK beam trawl survey and get more weight than the other fleets up to age 5 . The commercial fleets provide around $50 \%$ of the weight of ages 4 and older. The contribution of Fshrinkage is minor for all ages. Fishing mortalities and population numbers estimated from the final run are given in Tables 8.2.10 and 8.2.11, and summarized in Table 8.2.12. The 2006 and 2007 above average year classes have led to an increase in SSB in 2009. The 2008 year class appears to be weak. However in last year's assessment, the 2007 year class was estimated to be weak but is now being estimated to be above average.

Retrospective analysis (Figure 8.2.7) was run without the short UK (E\&W) FSP tun-ing-series, and indicates a strong downward revision of the 2001 year class strength, going from the second strongest year class in history to a value much closer to longterm GM. The changes to the assessment made at WKFLAT 2010 have arguably re-
solved the retrospective bias seen in last previous assessments where there was a sequential downwards revision of F and a commensurate revision in SSB. This assessment shows that retrospective pattern in F is substantially reduced and the F level remains relatively stable throughout the time-series.

## Comparison with previous assessments

Fishing mortality has decreased in 2009 (0.44) and SSB is estimated to have increased to 1833 t . Last year, fishing mortality and SSB in 2008 were estimated to be 0.64 and 1500 t ; this year's estimates for 2008 are 0.71 and 1653 t , an upward revision of $11 \%$ and $10 \%$ respectively. It should be noted that the $\mathrm{F}(\mathrm{bar})$ age range was revised at WKFLAT in 2010 and is now for ages 3 to 6 .

There is now no bias in the retrospective analysis. Historical stock trends are strongly converged. The most recent estimates of F show a slight underestimation with a slight overestimation in SSB.

## State of the stock

A summary of the final assessment is given in Table 8.2.12 and Figure 8.2.8. Spawn-ing-stock biomass (SSB) was stable during the period 1981-1987, peaked above 5000 t during 1988-1990 following good recruitments in the mid-1980s, and then decreased to around 2400 t in 1995-1996. Since then SSB increased following the good 1996 year class but has subsequently declined steadily to the lowest level in the time-series of around 1650 t in 2008. The SSB estimate for 2009 shows a slight increase from this level.

Fishing mortality showed a gradually increasing trend up until the mid 1990s, then a slight decline followed by a sharp increase up to 2007. This assessment shows a reduction in F in 2008 followed by a larger reduction again in 2009. The decline in F in 2009 is evidenced by a large reduction in effort observed for the UK beam trawl survey and a corresponding reduction in Belgian beam trawl effort although the decline in F may be overestimated.

Two periods of below average recruitments in the period 1989-1994 and from 19982006 have contributed to the decrease in yield and SSB. This assessment estimates that only two year classes have been above the long-term GM80-07 (5981) since 2000.

### 8.2.3 Short-term projections

In recent years, no catch forecast has been provided by the Working Group due to the persistent strong bias in the estimation of $F$ in the most recent years, the degree of which was unpredictable.

Given the changes made to the assessment at the Benchmark meeting in February 2010, this bias was removed making it now suitable to provide more detailed management advice. This year's forecast was run with F scaled to the last year due to the large fall in F observed in the final year of the assessment.

## Estimating year-class abundance

The 2007 year class is now estimated at 5.7 million at age 1 , which is over four times the estimate from last year's assessment ( 1.4 million) which was the lowest value in the time-series. This year's estimate shows that UK-WEC-BTS survey takes $57 \%$ of the weight, the UK (E\&W) FSP taking 42\% of the weight and the remainder coming from F shrinkage.

The 2008 year class is estimated to be around 2.0 million with $92 \%$ of the weight coming from the UK-WEC-BTS. This is the lowest value in the time-series but given the revised estimate of the 2007 year class, this should be considered to be highly uncertain.

Working Group estimates of year-class strength used for prediction can be summarised as follows:

Recruitment-at-age 1.

| Year class | Thousands | Basis | Surveys | Commercial | Shrinkage |
| :---: | :---: | :--- | :---: | :---: | :---: | :---: |
| 2007 | 5686 | XSA | $98 \%$ | - | $\mathbf{2 \%}$ |
| 2008 | 1962 | XSA | $92 \%$ | - | $8 \%$ |
| 2009 | 4969 | GM $(89-07)$ | - | - | - |
| 2010 | 4969 | GM $(89-07)$ | - | - | - |

The input values for the catch forecast (using the MFDP software) are given in Table 8.2.13. The F at age values used were calculated as the mean of the XSA values from 2007-2009, scaled to the final year. Catch and stock weights-at-age were also the mean of the period 2007-2009. Stock numbers-at-age in 2010 for ages 2 and older were obtained from the XSA. SSB values are calculated for 1 January.

Table 8.2.14 gives the management option table from the status quo catch prediction, and short-term results are shown in Figure 8.2.10. Assuming status quo F (Fsq = 0.44) implies landings of 1079 t in 2010 and 1040 t in 2011. (The TAC for 2010 is 4274 t . for VIId,e). SSB is predicted to rise from 2165 t in 2010 to 2337 t in 2011 and 2439 t in 2011. These results are discussed further in Section 8.2.10.

The detailed output for the status quo F forecast by age group is given in Table 8.2.15, and the estimated contributions of recent year classes to the predicted catches and SSBs are given in Table 8.2.16. The assumptions of $\mathrm{GM}_{1989-07}$ recruitment are predicted to contribute $13 \%$ to the landings in 2011 and $33 \%$ to SSB in 2012.

The stock and recruitment scatter plot is given in Figure 8.2.11.

### 8.2.4 FMSY evaluation

To derive an Fmsy estimate the SRMSYMC package was employed and $\mathrm{F}_{\text {msy }}$ was calculated based on the three common stock-recruit relationships; Ricker, Beverton-Holt and smooth Hockey stick. Figure 8.2.12 illustrates the curves and the percentiles of estimates with converged $\mathrm{F}_{\mathrm{mSY}}$ values for the three models estimated by the package. Models were fitted using 1000 MCMC re-samples. For all three stock-recruit relationships (SRR), all re-samples allowed FmSy and Fcrash values to be determined. All three models show that there is little evidence of a stock-recruitment relationship with only limited information as to the trends at extreme levels of SSB.

The smooth hockey stick model showed a 'break-off' point in the SRR that was inconsistent with the data and as such was rejected. The yield-per-recruit estimates were highly uncertain with high CV's. Therefore these estimates were also rejected. The yield-per-recruit output from the model is shown in Figure 8.2.15.

Figures 8.2.13-8.2.14 show box plots of $\mathrm{F}_{\mathrm{msy}}$ and $\mathrm{F}_{\text {crash }}$ together with values of $\mathrm{F}_{\mathrm{pa}}$ and F2009 ( $\mathrm{F}_{\text {lim }}$ is not defined) for the Ricker and the Beverton-Holt SRR models. The two SRR models have very different levels of estimated Fmš.

The full diagnostics for all model fits are shown at Table 8.2.18 and Figure 8.2.16.

Therefore, the suggested level of Fmsy for this stock is F's within the range of 0.14 and 0.31 .

| Stock-Recruit relationship Model |  |  |  | FMSY | FCrash |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Ricker | 0.312 | 0.750 |  |  |  |
| Beverton-Holt | 0.143 | 0.781 |  |  |  |

### 8.2.5 Biological reference points

ICES previously defined the following precautionary reference points.

| Flim | Not defined | $\mathrm{F}_{\mathrm{pa}}$ | 0.45 | (low probability that <br> SSBMT<Ba) |
| :--- | :--- | :--- | :--- | :--- |
| $\mathrm{B}_{\text {lim }}$ | 1300 t (equal to $\left.\mathrm{B}_{\mathrm{loss}}\right)$ | $\mathrm{B}_{\mathrm{pa}}$ | 2500 t | (equal to $\left.\mathrm{M}_{\text {BaL }}\right)$ |

However the Working Groups since 2004 had considered the biological reference points for this stock as unreliable for the following reasons:

- The stock-recruitment relation shows no evidence of reduced recruitment at low stock levels;
- The basis for $\mathrm{B}_{\mathrm{pa}}$ is weak, and heavily dependent on two consecutive points (1985 and 1986);
- $\mathrm{F}_{\mathrm{pa}}$ is based on $\mathrm{B}_{\mathrm{pa}}$, and then this reference point is also rejected;
- WKFLAT 2010 examined the stock dynamics provided by the migration model to determine appropriate biological reference points for this stock on the basis of the new assessment. It concluded that the historical reference points for this stock were no longer appropriate as the new assessment indicated significant changes to the historical perspective of the stock caused by the inclusion of catches from VIId in the VIIe plaice stock. WKFLAT 2010 provided a number of options for the preferred assessment methodology none of which are entirely satisfactory, but suggested that a Btrigger could reasonably be set at 2200 t provided that the move towards a suitable proxy of $\mathrm{F}_{\text {msy }}$ is effective to avoid further deterioration of SSB. Fmsy for plaice needs to consider the management target set for sole 7e as plaice are taken largely as a bycatch in the same fisheries, and because there is a currently accepted management target of $\mathrm{F}=0.27$ for sole VIIe.

The current assessment with the VIId migration correction shows a 'scaled-up' timeseries of SSB compared to last year's assessment. This shows that SSB has been below $2500 \mathrm{t}\left(\mathrm{B}_{\mathrm{pa}}\right)$ since 2003. Increases to SSB have been observed since then and recruitment does not appear to have been limited. F has been between 0.55 and 0.75 for almost the entire time-series, well above $\mathrm{F}_{\mathrm{pa}}$, without apparent stock collapse.

## Yield-per-recruit analysis

Results for the deterministic yield and SSB per recruit (using program MFYPR), conditional on the recent exploitation pattern, are given in Table 8.2.17 and Figure 8.2.10. $\mathrm{F}_{\max }$ is given by a reference F of 0.27 , around $62 \%$ of $\mathrm{F}_{\text {sq. }}$. Long-term yield and SSB (at $\mathrm{F}_{\text {sq }}$ and assuming $\mathrm{GM}_{89-07}$ recruitment $=4.969$ million) are given as 1450 t and 3220 t respectively.

### 8.2.6 Management plans

There is no management plan in place for this stock.

### 8.2.7 Uncertainties and bias in assessment and forecast

The WKFLAT (2010) Benchmark Assessment reviewed the stock identity, data and assessment model suggesting modifications to resolve long-term problems with the assessment, particularly the retrospective bias. As a result, the retrospective bias has been reduced and provision of a short-term forecast is now possible. The revision introduced new uncertainties into a portion of the data ( $\sim 10 \%$ ). A spawning migration correction assumes that a constant $15 \%$ of quarter 1 catches in VIId to originate from VIIe, based on historical tagging information. This proportion makes no provision for changes in the relative sizes of the two populations. In addition, this correction utilises the age structure of the VIId catches, representing a mix of age structure from VIIe, VIId and portions of the Area IV populations migrating into VIId for spawning.

There is a heavy reliance on the age composition data derived from UK (E\&W) sample data. Around $20 \%$ of the landings for this stock come from countries that do not provide age based data and this situation is improved only slightly once the migration correction data from VIId is added in. Survivor estimates for ages 1 and 2 almost entirely come from the UK survey data and some consideration should be given to using age 2 information from the commercial tuning fleets.

UK Discard data indicate low discard levels in the second half of the year, and overall that discarding for this stock is variable but relatively low compared to other plaice stocks. As the time-series of data expands, the WG will be able to better determine how to include this data in the assessment appropriately.

Both the UK-WEC_BTS and the UK (E\&W) FSP surveys are spatially restricted to the same area as the commercial tuning fleets and little information exists on stock dynamics on the French coast.

### 8.2.8 Recommendation for next Benchmark

A benchmark assessment was carried out for this stock in February 2010.

### 8.2.9 Management considerations

The assessment model developed at WKFLAT 2010 includes an element of catch and associated age based data from the adjacent ICES plaice stock in VIId. Therefore a reciprocal removal of this data must occur with that stock. This adjustment is made to account for the spawning migration that occurs between the two areas. However, WKFLAT recognised that this is based on historical tagging information that assumes the historical patterns of migration has persisted and that the relative size of the subpopulations is roughly stable. WKFLAT suggested that tagging experiments should be reinitiated to provide a more up-to-date and precise estimate of the level of migration.

The stock unit (Division VIIe) does not correspond with the management unit (Divisions VIId and VIIe). This hampers effective management of plaice in the Western English Channel, but because components of the VIIe stock are also taken during spawning time in Area VIId, some provision must be made in management to accommodate effective management of both plaice stocks.

Plaice are taken as a bycatch in the beam trawl fishery mainly targeting sole, and as part of a mixed demersal fishery by otter trawlers. Therefore the restrictions under the management plan for sole should also benefit the plaice stocks. In addition to the days-at-sea regulations there has been a recent UK decommissioning scheme that has reduced the number of beam trawlers in the southwest fleet. Fishing mortality in 2009 is estimated to have declined heavily which is consistent with the decline observed in beam trawl effort in 2009.

The assessment is now able to accurately estimate recent trends in F and historical trends are estimated with some certainty. Fishing mortality is estimated to be well above long-term targets with some certainty.

Table 8.2.1 Plaice in VIIe. Nominal landings ( t ) in Division VIIe, as used by Working Group.

| Year | Belgium | Denmark | France | UK (Engl. \& Wales) | Others | Total reported | Unallocated ${ }^{1}$ | Total | VIIe stock caught in VIId $^{4}$ | As used by WG |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 5 | - 3 | 323 | 312 | - | 640 | - | 640 |  | 640 |
| 1977 | 3 | - 3 | 336 | 363 | - | 702 | - | 702 | - | 702 |
| 1978 | 3 | - 3 | 314 | 467 | - | 784 | - | 784 | - | 784 |
| 1979 | 2 | - 3 | 458 | 515 | - | 975 | 2 | 977 | - | 977 |
| 1980 | 23 | - 3 | 325 | 609 | 9 | 966 | 113 | 1079 | 136 | 1215 |
| 1981 | 27 | - | 537 | 953 | - | 1517 | -16 | 1501 | 245 | 1746 |
| 1982 | 81 | - | 363 | 1109 | - | 1553 | 135 | 1688 | 250 | 1938 |
| 1983 | 20 | - | 371 | 1195 | - | 1586 | -91 | 1495 | 259 | 1754 |
| 1984 | 24 | - | 278 | 1144 | - | 1446 | 101 | 1547 | 266 | 1813 |
| 1985 | 39 | - | 197 | 1122 | - | 1358 | 83 | 1441 | 310 | 1751 |
| 1986 | 26 | - | 276 | 1389 | -1 | 1691 | 119 | 1810 | 351 | 2161 |
| 1987 | 68 | - | 435 | 1419 | - | 1922 | 36 | 1958 | 430 | 2388 |
| 1988 | 90 | - | 584 | 1654 | - | 2328 | 130 | 2458 | 536 | 2994 |
| 1989 | 89 | - | 4481 | 1708 | 2 | 2247 | 111 | 2358 | 450 | 2808 |
| 1990 | 82 | 2 | N/A ${ }^{2}$ | 1885 | 18 | 1987 | 606 | 2593 | 465 | 3058 |
| 1991 | 57 | - | 2511 | 1323 | 16 | 1647 | 201 | 1848 | 402 | 2250 |
| 1992 | 25 | - | 419 | 1102 | 14 | 1560 | 64 | 1624 | 326 | 1950 |
| 1993 | 56 | - | 284 | 1080 | 24 | 1444 | -27 | 1417 | 274 | 1691 |
| 1994 | 10 | - | 277 | 998 | 3 | 1288 | -132 | 1156 | 315 | 1471 |
| 1995 | 13 | - | 288 | 857 | - | 1158 | -127 | 1031 | 264 | 1295 |
| 1996 | 4 | - | 279 | 855 | - | 1138 | -94 | 1044 | 277 | 1321 |
| 1997 | 6 | - | 329 | 1038 | 1 | 1374 | -51 | 1323 | 331 | 1654 |
| 1998 | 22 | - | 327 | 892 | 1 | 1242 | -111 | 1131 | 299 | 1430 |
| 1999 | 12 | - | 1941 | 947 | - | 1153 | 118 | 1271 | 345 | 1616 |
| 2000 | 4 | - | 360 | 926 | + | 1290 | -9 | 1281 | 397 | 1678 |
| 2001 | 12 | - | 303 | 797 | - | 1112 | -6 | 1106 | 273 | 1379 |
| 2002 | 27 | - | 242 | 978 | + | 1253 | 4 | 1257 | 351 | 1608 |
| 2003 | 39 | - | 216 | 985 | - | 1217 | 1 | 1218 | 260 | 1478 |
| 2004 | 46 | - | 184 | 912 | - | 1142 | 12 | 1154 | 248 | 1402 |
| 2005 | 48 | - | 198 | 887 | - | 1133 | 66 | 1199 | 171 | 1370 |
| 2006 | 52 | - | 223 | 966 | - | 1241 | 72 | 1313 | 153 | 1466 |
| 2007 | 84 | - | 201 | 677 | - | 962 | 41 | 1003 | 181 | 1184 |
| 2008 | 66 | - | 105 | 669 | - | 840 | 134 | 974 | 170 | 1144 |
| 2009 | 60 | - | - | 724 | 3 | 787 | 129 | 916 | 127 | 1043 |

[^18]Table 8.2.2 Division VIIe PLAICE effort and CPUE data
The UK (E\&W) data are for vessels > 12m and are corrected for fishing power (based on GRT). All effort data are in fishing hours, CPUE data are in $\mathrm{kg} / \mathrm{hr}$ for the commercial fleets and in $\mathrm{kg} / 10 \mathrm{~km}$ towed for the autumn beam trawl survey

| Year | $\begin{gathered} \text { (CPUE) } \\ (\mathrm{kg} / \mathrm{hr}) . \\ \hline \end{gathered}$ |  |  |  |  |  | Effort (000 hours) |  | Landings (tonnes) |  | $\begin{gathered} \text { (CPUE) } \\ (\mathrm{kg} / 10 \mathrm{~km}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | West Sector |  | North Sector |  | South Sector |  |  |  |  |  |  |
|  | Otter | Beam | Otter | Beam | Otter | Beam | Otter | Beam | Otter | Beam | Survey |
| 1972 | 2.31 | - | 4.50 | - | 0.00 | - | 64.60 | - | 194.36 | - | - |
| 1973 | 2.25 | - | 3.85 | - | 0.00 | - | 69.54 | - | 200.45 | - | - |
| 1974 | 1.65 | - | 3.47 | - | 2.94 | - | 50.09 | - | 121.03 | - | - |
| 1975 | 1.78 | - | 3.53 | - | 2.54 | - | 54.69 | - | 132.95 | - | - |
| 1976 | 1.89 | - | 3.62 | - | 4.14 | - | 56.13 | - | 144.56 | - | - |
| 1977 | 1.37 | - | 3.10 | - | 4.96 | - | 55.40 | - | 117.72 | - | - |
| 1978 | 1.61 | 5.41 | 3.63 | 10.35 | 4.24 | 11.84 | 48.80 | 22.09 | 114.02 | 204.69 | - |
| 1979 | 1.84 | 4.16 | 4.58 | 7.37 | 1.64 | 6.58 | 49.92 | 39.38 | 142.52 | 233.81 | - |
| 1980 | 2.02 | 3.15 | 5.82 | 6.06 | 0.67 | 6.45 | 49.95 | 62.16 | 150.69 | 335.16 | - |
| 1981 | 2.61 | 4.44 | 10.98 | 8.35 | 7.30 | 8.33 | 46.88 | 65.29 | 257.28 | 471.20 | - |
| 1982 | 3.28 | 4.43 | 10.77 | 9.23 | 0.00 | 7.69 | 38.51 | 81.59 | 249.60 | 611.52 | - |
| 1983 | 2.57 | 2.76 | 11.03 | 9.64 | 8.10 | 5.71 | 52.59 | 103.07 | 303.04 | 612.16 | - |
| 1984 | 2.95 | 4.08 | 10.92 | 10.38 | 2.43 | 7.80 | 52.89 | 87.63 | 281.94 | 575.22 | - |
| 1985 | 2.60 | 3.79 | 8.81 | 9.00 | 0.09 | 6.38 | 57.69 | 92.19 | 255.86 | 540.61 | 15.21 |
| 1986 | 3.25 | 6.30 | 10.94 | 12.21 | 10.17 | 6.85 | 49.52 | 76.33 | 315.08 | 602.07 | 16.46 |
| 1987 | 3.56 | 5.37 | 11.02 | 9.69 | 3.63 | 7.45 | 45.11 | 87.05 | 329.97 | 672.81 | 20.59 |
| 1988 | 3.90 | 3.50 | 15.38 | 6.51 | 5.04 | 4.85 | 53.40 | 103.36 | 433.20 | 564.72 | 25.34 |
| 1989 | 2.69 | 6.50 | 10.87 | 14.25 | 1.42 | 6.88 | 54.71 | 109.95 | 315.73 | 900.19 | 14.80 |
| 1990 | 2.95 | 6.52 | 7.77 | 15.64 | 3.55 | 10.17 | 53.05 | 100.95 | 268.81 | 990.05 | 11.60 |
| 1991 | 2.80 | 6.16 | 5.08 | 13.24 | 0.41 | 7.47 | 40.79 | 83.57 | 152.93 | 721.46 | 8.72 |
| 1992 | 1.92 | 6.30 | 3.51 | 10.61 | 3.06 | 9.69 | 39.91 | 80.87 | 105.41 | 695.70 | 7.45 |
| 1993 | 1.39 | 6.14 | 3.03 | 11.04 | 5.46 | 7.17 | 39.17 | 83.92 | 81.77 | 655.48 | 6.16 |
| 1994 | 1.46 | 4.62 | 2.48 | 9.17 | 2.11 | 6.47 | 38.77 | 100.42 | 63.67 | 650.99 | 5.70 |
| 1995 | 1.61 | 4.60 | 1.99 | 6.29 | 2.36 | 5.40 | 35.45 | 100.80 | 60.20 | 531.06 | 5.13 |
| 1996 | 2.00 | 3.09 | 2.49 | 6.66 | 11.62 | 4.39 | 30.54 | 116.45 | 64.83 | 482.18 | 5.97 |
| 1997 | 2.69 | 3.50 | 3.08 | 7.16 | 1.56 | 5.58 | 33.28 | 108.39 | 99.05 | 561.74 | 9.82 |
| 1998 | 1.65 | 2.97 | 4.13 | 6.10 | 1.85 | 3.03 | 29.80 | 111.17 | 73.30 | 459.22 | 8.74 |
| 1999 | 1.39 | 3.49 | 3.60 | 8.55 | 1.11 | 4.59 | 27.52 | 103.56 | 59.67 | 576.76 | 8.42 |
| 2000 | 0.81 | 2.98 | 4.00 | 6.63 | 1.25 | 3.72 | 30.49 | 118.83 | 61.82 | 541.33 | 11.31 |
| 2001 | 0.89 | 2.30 | 3.03 | 5.45 | 3.14 | 3.61 | 31.90 | 143.27 | 48.82 | 527.38 | 10.56 |
| 2002 | 0.90 | 2.90 | 4.18 | 6.52 | 0.56 | 3.45 | 28.35 | 139.83 | 57.44 | 651.04 | 8.05 |
| 2003 | 0.96 | 3.26 | 2.10 | 8.18 | 0.50 | 2.89 | 25.06 | 159.95 | 36.88 | 743.07 | 7.96 |
| 2004 | 0.88 | 3.38 | 2.01 | 6.16 | 0.19 | 2.80 | 25.58 | 158.68 | 37.98 | 701.17 | 4.53 |
| 2005 | 0.88 | 2.62 | 2.13 | 8.20 | 3.48 | 2.75 | 21.13 | 157.81 | 29.44 | 691.27 | 7.02 |
| 2006 | 0.96 | 2.68 | 3.41 | 6.97 | 1.71 | 2.50 | 21.06 | 161.44 | 28.57 | 665.16 | 7.47 |
| 2007 | 0.68 | 1.71 | 1.95 | 4.55 | 1.31 | 2.13 | 22.35 | 158.01 | 27.27 | 472.27 | 7.94 |
| 2008 | 0.94 | 1.83 | 2.07 | 4.88 | 0.71 | 2.06 | 19.86 | 158.50 | 25.72 | 465.09 | 8.18 |
| 2009 | 1.26 | 2.62 | 2.23 | 7.59 | 1.78 | 3.48 | 21.40 | 122.53 | 32.45 | 521.17 | 12.85 |

Table 8.2.3. Plaice in VIIe. Annual length distribution by fleet (2009)

| Length (cm) | UK (England \& Wales) |  |  |
| :---: | :---: | :---: | :---: |
|  |  |  | All gears (excl. beam) |
|  | 23 | 43 | 112 |
|  | 24 | 0 | 544 |
|  | 25 | 3142 | 2694 |
|  | 26 | 22054 | 12196 |
|  | 27 | 47104 | 25147 |
|  | 28 | 77147 | 47446 |
|  | 29 | 101185 | 61758 |
|  | 30 | 105559 | 61043 |
|  | 31 | 125512 | 61622 |
|  | 32 | 147176 | 51168 |
|  | 33 | 124418 | 44881 |
|  | 34 | 107038 | 38625 |
|  | 35 | 87257 | 27375 |
|  | 36 | 65046 | 21741 |
|  | 37 | 45314 | 16069 |
|  | 38 | 38344 | 12542 |
|  | 39 | 25152 | 8871 |
|  | 40 | 22212 | 5664 |
|  | 41 | 16959 | 5955 |
|  | 42 | 14118 | 3022 |
|  | 43 | 9052 | 2531 |
|  | 44 | 7910 | 2210 |
|  | 45 | 7402 | 2112 |
|  | 46 | 5177 | 1199 |
|  | 47 | 4149 | 1039 |
|  | 48 | 3442 | 1042 |
|  | 49 | 2318 | 775 |
|  | 50 | 1851 | 572 |
|  | 51 | 2133 | 684 |
|  | 52 | 1788 | 134 |
|  | 53 | 1175 | 278 |
|  | 54 | 1125 | 424 |
|  | 55 | 1399 | 223 |
|  | 56 | 901 | 119 |
|  | 57 | 344 | 150 |
|  | 58 | 703 | 205 |
|  | 59 | 557 | 93 |
|  | 60 | 149 | 78 |
|  | 61 | 42 | 55 |
|  | 62 | 0 | 5 |
|  | 63 | 155 | 0 |
|  | 64 | 0 | 236 |
|  | 65 | 11 |  |
|  | 66 | 0 |  |
|  | 67 | 21 |  |
|  | 68 | 8 |  |
| Total |  | 1226591 | 522639 |

Table 8.2.4 Plaice in VIle. Catch numbers-at-age.

| Table 1 | Catch numbers at age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 19 | 41 | 72 | 3 | 77 | 3 | 10 | 74 | 12 | 10 |
| 2 | 814 | 723 | 310 | 790 | 970 | 727 | 1025 | 1258 | 1932 | 352 |
| 3 | 800 | 2268 | 2131 | 893 | 1864 | 1605 | 2532 | 2303 | 5179 | 2960 |
| 4 | 252 | 591 | 1420 | 1702 | 702 | 1399 | 963 | 1407 | 1160 | 3014 |
| 5 | 230 | 120 | 263 | 593 | 531 | 157 | 488 | 657 | 464 | 843 |
| 6 | 62 | 103 | 89 | 104 | 197 | 255 | 116 | 233 | 155 | 274 |
| 7 | 63 | 21 | 83 | 41 | 92 | 142 | 129 | 90 | 116 | 121 |
| 8 | 23 | 47 | 17 | 50 | 30 | 28 | 68 | 52 | 40 | 97 |
| 9 | 13 | 19 | 28 | 2 | 33 | 16 | 29 | 45 | 25 | 32 |
| +gp | 138 | 95 | 122 | 100 | 51 | 52 | 62 | 52 | 53 | 101 |
| TOTALNUM | 2415 | 4027 | 4534 | 4276 | 4546 | 4383 | 5421 | 6170 | 9136 | 7805 |
| TONSLAND | 1215 | 1746 | 1938 | 1754 | 1813 | 1751 | 2161 | 2388 | 2994 | 2808 |
| SOPCOF\% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 57 | 41 | 90 | 36 | 84 | 6 | 15 | 7 | 7 | 19 |
| 2 | 391 | 691 | 841 | 844 | 409 | 421 | 1160 | 963 | 636 | 678 |
| 3 | 3408 | 1352 | 1430 | 1488 | 1707 | 818 | 774 | 2443 | 1732 | 2480 |
| 4 | 2757 | 1943 | 760 | 650 | 878 | 986 | 403 | 486 | 1158 | 1219 |
| 5 | 1222 | 973 | 654 | 266 | 256 | 269 | 392 | 185 | 159 | 414 |
| 6 | 272 | 528 | 452 | 272 | 111 | 120 | 127 | 155 | 66 | 94 |
| 7 | 135 | 106 | 264 | 219 | 119 | 58 | 60 | 80 | 61 | 38 |
| 8 | 80 | 46 | 72 | 171 | 83 | 84 | 41 | 34 | 23 | 40 |
| 9 | 57 | 33 | 33 | 40 | 86 | 69 | 48 | 18 | 21 | 17 |
| +gp | 73 | 51 | 50 | 86 | 65 | 90 | 107 | 101 | 63 | 46 |
| TOTALNUM | 8451 | 5764 | 4646 | 4071 | 3797 | 2920 | 3127 | 4472 | 3926 | 5046 |
| TONSLAND | 3058 | 2250 | 1950 | 1691 | 1471 | 1295 | 1321 | 1654 | 1430 | 1616 |
| SOPCOF\% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 46 | 6 | 188 | 23 | 21 | 22 | 18 | 3 | 5 | 5 |
| 2 | 399 | 585 | 1400 | 1004 | 600 | 831 | 1089 | 428 | 1015 | 735 |
| 3 | 1331 | 946 | 1251 | 1208 | 1644 | 1034 | 1448 | 1168 | 781 | 1328 |
| 4 | 2069 | 795 | 597 | 622 | 600 | 858 | 543 | 723 | 563 | 288 |
| 5 | 496 | 950 | 428 | 207 | 349 | 282 | 388 | 287 | 252 | 143 |
| 6 | 181 | 145 | 511 | 172 | 102 | 146 | 121 | 196 | 107 | 74 |
| 7 | 38 | 79 | 116 | 224 | 75 | 52 | 60 | 70 | 83 | 29 |
| 8 | 14 | 19 | 49 | 54 | 96 | 50 | 29 | 30 | 32 | 20 |
| 9 | 22 | 12 | 13 | 41 | 44 | 53 | 22 | 10 | 15 | 7 |
| +gp | 52 | 37 | 42 | 39 | 38 | 44 | 45 | 49 | 28 | 16 |
| TOTALNUM | 4648 | 3574 | 4595 | 3594 | 3569 | 3372 | 3764 | 2962 | 2882 | 2646 |
| TONSLAND | 1678 | 1379 | 1608 | 1478 | 1402 | 1370 | 1466 | 1184 | 1144 | 1043 |
| SOPCOF\% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table 8.2.5 Plaice in VIIe. Catch weights-at-age.
Table 2 Catch weights at age (kg)

| 1989 |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAR |  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.248 | 0.144 | 0.186 | 0.106 | 0.136 | 0.098 | 0.171 | 0.252 | 0.134 | 0.156 |
|  | 2 | 0.337 | 0.268 | 0.273 | 0.221 | 0.238 | 0.214 | 0.257 | 0.288 | 0.215 | 0.217 |
|  | 3 | 0.428 | 0.389 | 0.36 | 0.33 | 0.343 | 0.328 | 0.346 | 0.337 | 0.303 | 0.285 |
|  | 4 | 0.519 | 0.507 | 0.447 | 0.432 | 0.447 | 0.437 | 0.438 | 0.403 | 0.399 | 0.36 |
|  | 5 | 0.612 | 0.622 | 0.532 | 0.529 | 0.55 | 0.543 | 0.533 | 0.48 | 0.504 | 0.44 |
|  | 6 | 0.706 | 0.733 | 0.619 | 0.617 | 0.654 | 0.644 | 0.632 | 0.572 | 0.618 | 0.528 |
|  | 7 | 0.801 | 0.841 | 0.702 | 0.699 | 0.757 | 0.743 | 0.734 | 0.679 | 0.74 | 0.622 |
|  | 8 | 0.898 | 0.946 | 0.786 | 0.775 | 0.861 | 0.837 | 0.84 | 0.799 | 0.87 | 0.723 |
|  | 9 | 0.996 | 1.047 | 0.869 | 0.844 | 0.965 | 0.928 | 0.95 | 0.933 | 1.009 | 0.83 |
|  | + gp | 1.404 | 1.387 | 1.217 | 1.027 | 1.39 | 1.253 | 1.427 | 1.388 | 1.357 | 1.122 |
| SOPCOFAC | 0.9999 | 1.0007 | 0.9999 | 1.0003 | 1.0000 | 0.9996 | 0.9993 | 0.9997 | 0.9991 | 1.0001 |  |


| 1999 |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAR |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.236 | 0.194 | 0.242 | 0.212 | 0.201 | 0.213 | 0.173 | 0.188 | 0.179 | 0.107 |
|  | 2 | 0.267 | 0.245 | 0.282 | 0.269 | 0.258 | 0.281 | 0.266 | 0.259 | 0.239 | 0.196 |
|  | 3 | 0.308 | 0.306 | 0.335 | 0.332 | 0.322 | 0.353 | 0.36 | 0.334 | 0.294 | 0.282 |
|  | 4 | 0.359 | 0.377 | 0.401 | 0.405 | 0.391 | 0.429 | 0.455 | 0.412 | 0.411 | 0.364 |
|  | 5 | 0.421 | 0.456 | 0.481 | 0.484 | 0.464 | 0.507 | 0.551 | 0.494 | 0.526 | 0.444 |
|  | 6 | 0.493 | 0.545 | 0.574 | 0.571 | 0.543 | 0.588 | 0.647 | 0.58 | 0.638 | 0.521 |
|  | 7 | 0.577 | 0.643 | 0.68 | 0.667 | 0.628 | 0.674 | 0.743 | 0.669 | 0.747 | 0.596 |
|  | 8 | 0.67 | 0.75 | 0.799 | 0.769 | 0.717 | 0.763 | 0.84 | 0.762 | 0.853 | 0.667 |
|  | 9 | 0.775 | 0.866 | 0.933 | 0.88 | 0.812 | 0.855 | 0.938 | 0.86 | 0.958 | 0.735 |
|  | +gp | 1.078 | 1.221 | 1.317 | 1.202 | 1.117 | 1.055 | 1.17 | 1.11 | 1.274 | 0.95 |
| SOPCOFAC | 0.9996 | 1.0004 | 0.9996 | 1.0000 | 1.0002 | 0.9998 | 1.0006 | 0.9992 | 1.0004 | 1.0000 |  |


| YEAR |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.117 | 0.167 | 0.193 | 0.147 | 0.254 | 0.226 | 0.206 | 0.186 | 0.208 | 0.096 |
|  | 2 | 0.204 | 0.231 | 0.246 | 0.25 | 0.293 | 0.287 | 0.276 | 0.259 | 0.279 | 0.238 |
|  | 3 | 0.29 | 0.305 | 0.306 | 0.352 | 0.342 | 0.354 | 0.352 | 0.334 | 0.356 | 0.376 |
|  | 4 | 0.375 | 0.384 | 0.372 | 0.45 | 0.4 | 0.426 | 0.434 | 0.412 | 0.438 | 0.509 |
|  | 5 | 0.459 | 0.468 | 0.446 | 0.548 | 0.468 | 0.504 | 0.521 | 0.493 | 0.526 | 0.637 |
|  | 6 | 0.542 | 0.558 | 0.525 | 0.641 | 0.545 | 0.586 | 0.614 | 0.577 | 0.619 | 0.761 |
|  | 7 | 0.624 | 0.654 | 0.612 | 0.734 | 0.632 | 0.674 | 0.712 | 0.663 | 0.718 | 0.88 |
|  | 8 | 0.705 | 0.754 | 0.706 | 0.822 | 0.728 | 0.766 | 0.814 | 0.752 | 0.822 | 0.995 |
|  | 9 | 0.784 | 0.861 | 0.806 | 0.91 | 0.833 | 0.864 | 0.923 | 0.844 | 0.932 | 1.105 |
|  | + gp | 1.029 | 1.272 | 1.137 | 1.231 | 1.189 | 1.106 | 1.165 | 1.095 | 1.27 | 1.347 |
| SOPCOFAC | 0.9997 | 1.0001 | 0.9998 | 1.0003 | 1.0005 | 1.0002 | 1.0003 | 1.0001 | 1.0002 | 0.9994 |  |

Table 8.2.6 Plaice in VIle. Stock weights-at-age.
Table 3 Stock weights at age (kg)

| YEAR |  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.1140 | 0.1260 | 0.1080 | 0.1160 | 0.1110 | 0.1120 | 0.0960 | 0.0680 | 0.1030 | 0.1380 |
|  | 2 | 0.2270 | 0.2500 | 0.2140 | 0.2280 | 0.2220 | 0.2220 | 0.1950 | 0.1450 | 0.1840 | 0.2000 |
|  | 3 | 0.3380 | 0.3730 | 0.3180 | 0.3350 | 0.3340 | 0.3310 | 0.2970 | 0.2320 | 0.2750 | 0.2700 |
|  | 4 | 0.4470 | 0.4920 | 0.4190 | 0.4360 | 0.4460 | 0.4380 | 0.4010 | 0.3260 | 0.3730 | 0.3470 |
|  | 5 | 0.5540 | 0.6090 | 0.5170 | 0.5320 | 0.5600 | 0.5430 | 0.5070 | 0.4290 | 0.4810 | 0.4310 |
|  | 6 | 0.6600 | 0.7250 | 0.6150 | 0.6230 | 0.6730 | 0.6470 | 0.6150 | 0.5390 | 0.5980 | 0.5220 |
|  | 7 | 0.7640 | 0.8380 | 0.7100 | 0.7100 | 0.7880 | 0.7490 | 0.7270 | 0.6590 | 0.7230 | 0.6200 |
|  | 8 | 0.8670 | 0.9490 | 0.8020 | 0.7910 | 0.9030 | 0.8490 | 0.8400 | 0.7880 | 0.8580 | 0.7250 |
|  | 9 | 0.9670 | 1.0570 | 0.8930 | 0.8670 | 1.0180 | 0.9480 | 0.9550 | 0.9240 | 1.0020 | 0.8370 |
|  | +gp | 1.3510 | 1.4350 | 1.2550 | 1.0940 | 1.4980 | 1.3290 | 1.4420 | 1.3470 | 1.3630 | 1.1430 |


| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 1 | 0.2360 | 0.1820 | 0.2350 | 0.1880 | 0.1880 | 0.1910 | 0.1340 | 0.1710 | 0.1690 | 0.0690 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 0.2620 | 0.2320 | 0.2690 | 0.2410 | 0.2480 | 0.2620 | 0.2330 | 0.2480 | 0.2250 | 0.1710 |
| 3 | 0.3000 | 0.2920 | 0.3170 | 0.3020 | 0.3140 | 0.3360 | 0.3330 | 0.3290 | 0.2540 | 0.2700 |
| 4 | 0.3490 | 0.3620 | 0.3780 | 0.3710 | 0.3850 | 0.4130 | 0.4340 | 0.4140 | 0.3820 | 0.3650 |
| 5 | 0.4080 | 0.4420 | 0.4540 | 0.4470 | 0.4620 | 0.4950 | 0.5350 | 0.5030 | 0.5070 | 0.4570 |
| 6 | 0.4790 | 0.5310 | 0.5430 | 0.5310 | 0.5450 | 0.5800 | 0.6370 | 0.5960 | 0.6290 | 0.5450 |
| 7 | 0.5610 | 0.6310 | 0.6460 | 0.6230 | 0.6330 | 0.6680 | 0.7390 | 0.6940 | 0.7490 | 0.6310 |
| 8 | 0.6540 | 0.7400 | 0.7630 | 0.7230 | 0.7280 | 0.7600 | 0.8420 | 0.7950 | 0.8660 | 0.7120 |
| 9 | 0.7580 | 0.8580 | 0.8930 | 0.8300 | 0.8280 | 0.8560 | 0.9450 | 0.9010 | 0.9800 | 0.7910 |
| +gp | 1.0640 | 1.2230 | 1.2740 | 1.1450 | 1.1500 | 1.0640 | 1.1910 | 1.1760 | 1.3260 | 1.0400 |


| YEAR |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 0.0820 | 0.1390 | 0.1800 | 0.1000 | 0.2460 | 0.2050 | 0.1770 | 0.1560 | 0.1750 | 0.0230 |
|  | 2 | 0.1810 | 0.2040 | 0.2330 | 0.2110 | 0.2820 | 0.2660 | 0.2480 | 0.2290 | 0.2430 | 0.1670 |
|  | 3 | 0.2790 | 0.2770 | 0.2930 | 0.3190 | 0.3270 | 0.3340 | 0.3230 | 0.3050 | 0.3170 | 0.3080 |
|  | 4 | 0.3760 | 0.3560 | 0.3600 | 0.4250 | 0.3830 | 0.4060 | 0.4050 | 0.3850 | 0.3960 | 0.4430 |
|  | 5 | 0.4720 | 0.4410 | 0.4350 | 0.5290 | 0.4480 | 0.4840 | 0.4920 | 0.4670 | 0.4810 | 0.5740 |
| 6 | 0.5670 | 0.5310 | 0.5160 | 0.6300 | 0.5230 | 0.5670 | 0.5840 | 0.5510 | 0.5720 | 0.7000 |  |
|  | 7 | 0.6600 | 0.6270 | 0.6050 | 0.7280 | 0.6080 | 0.6560 | 0.6820 | 0.6390 | 0.6680 | 0.8210 |
|  | 8 | 0.7520 | 0.7290 | 0.7010 | 0.8240 | 0.7020 | 0.7490 | 0.7860 | 0.7300 | 0.7690 | 0.9380 |
| 9 | 0.8420 | 0.8360 | 0.8050 | 0.9180 | 0.8070 | 0.8490 | 0.8950 | 0.8230 | 0.8760 | 1.0500 |  |
|  | +gp | 1.1220 | 1.2530 | 1.1480 | 1.2630 | 1.1600 | 1.0950 | 1.1390 | 1.0780 | 1.2070 | 1.2980 |

Table 8.2.7 UK-WECBTS effort standardised plaice abundance indices

| age <br> year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1 9 8 5}$ | 0.00 | 82.16 | 75.37 | 72.36 | 113.06 | 20.35 | 15.83 | 8.29 | 0.75 | 0.00 | 2.26 |
| $\mathbf{1 9 8 6}$ | 0.00 | 61.62 | 86.67 | 168.60 | 64.33 | 23.70 | 2.71 | 12.19 | 1.35 | 0.00 | 1.35 |
| $\mathbf{1 9 8 7}$ | 0.74 | 398.98 | 110.17 | 104.21 | 54.34 | 27.54 | 21.59 | 10.42 | 5.95 | 5.95 | 2.98 |
| $\mathbf{1 9 8 8}$ | 0.00 | 108.40 | 289.33 | 265.15 | 75.65 | 17.16 | 8.58 | 7.80 | 3.12 | 4.68 | 3.12 |
| $\mathbf{1 9 8 9}$ | 0.00 | 18.71 | 42.26 | 169.63 | 113.49 | 13.88 | 6.64 | 8.45 | 4.83 | 3.62 | 10.87 |
| $\mathbf{1 9 9 0}$ | 0.00 | 14.23 | 21.63 | 125.24 | 49.53 | 42.70 | 1.14 | 3.42 | 0.57 | 3.42 | 3.98 |
| $\mathbf{1 9 9 1}$ | 1.16 | 12.81 | 15.73 | 36.70 | 46.02 | 36.11 | 23.88 | 5.24 | 0.00 | 0.58 | 1.75 |
| $\mathbf{1 9 9 2}$ | 0.00 | 77.31 | 22.38 | 36.62 | 12.21 | 20.35 | 10.17 | 8.65 | 1.53 | 2.54 | 2.03 |
| $\mathbf{1 9 9 3}$ | 0.00 | 11.10 | 37.00 | 31.71 | 12.69 | 6.87 | 13.21 | 6.87 | 5.81 | 1.06 | 1.06 |
| $\mathbf{1 9 9 4}$ | 0.00 | 16.52 | 15.54 | 47.60 | 14.57 | 4.86 | 0.97 | 4.37 | 6.31 | 3.89 | 0.97 |
| $\mathbf{1 9 9 5}$ | 0.00 | 26.72 | 24.58 | 24.04 | 25.65 | 6.41 | 2.14 | 2.67 | 3.21 | 0.53 | 2.14 |
| $\mathbf{1 9 9 6}$ | 0.54 | 17.90 | 57.49 | 16.27 | 9.22 | 13.56 | 2.71 | 0.54 | 1.63 | 3.80 | 4.34 |
| $\mathbf{1 9 9 7}$ | 0.00 | 28.69 | 66.04 | 106.63 | 12.99 | 3.25 | 6.50 | 3.79 | 0.54 | 0.54 | 3.79 |
| $\mathbf{1 9 9 8}$ | 0.00 | 43.67 | 67.39 | 67.39 | 45.83 | 4.85 | 3.23 | 3.77 | 2.16 | 0.00 | 1.62 |
| $\mathbf{1 9 9 9}$ | 0.53 | 20.22 | 23.42 | 96.86 | 28.21 | 15.97 | 1.60 | 1.06 | 3.19 | 2.13 | 1.06 |
| $\mathbf{2 0 0 0}$ | 0.00 | 26.57 | 34.79 | 69.51 | 99.00 | 21.13 | 12.30 | 0.60 | 1.11 | 0.00 | 2.77 |
| $\mathbf{2 0 0 1}$ | 11.52 | 17.91 | 35.78 | 28.65 | 62.57 | 54.75 | 13.79 | 7.08 | 0.00 | 1.69 | 2.81 |
| $\mathbf{2 0 0 2}$ | 0.00 | 76.78 | 56.50 | 48.17 | 12.91 | 13.06 | 22.18 | 2.97 | 1.11 | 0.00 | 1.11 |
| $\mathbf{2 0 0 3}$ | 0.00 | 15.82 | 75.35 | 32.84 | 27.52 | 2.47 | 9.91 | 14.86 | 3.96 | 0.00 | 1.10 |
| $\mathbf{2 0 0 4}$ | 0.00 | 6.71 | 19.82 | 35.67 | 14.03 | 6.10 | 1.83 | 0.61 | 6.10 | 0.00 | 2.44 |
| $\mathbf{2 0 0 5}$ | 0.80 | 16.31 | 40.42 | 48.71 | 37.42 | 6.90 | 1.71 | 1.43 | 2.81 | 1.18 | 1.47 |
| $\mathbf{2 0 0 6}$ | 0.00 | 29.77 | 55.43 | 55.78 | 16.45 | 16.89 | 1.44 | 2.06 | 0.00 | 2.44 | 1.08 |
| $\mathbf{2 0 0 7}$ | 0.00 | 20.44 | 50.35 | 66.58 | 18.67 | 14.93 | 3.31 | 3.04 | 0.28 | 1.38 | 2.21 |
| $\mathbf{2 0 0 8}$ | 0.00 | 8.54 | 83.46 | 38.71 | 17.67 | 6.87 | 4.48 | 5.44 | 2.00 | 0.57 | 1.72 |
| $\mathbf{2 0 0 9}$ | 1.74 | 9.40 | 90.88 | 124.18 | 16.93 | 8.50 | 6.36 | 4.65 | 2.68 | 0.58 | 1.45 |

Table 8.2.8 Plaice in VIle. Tuning fleet data available
W.ChanNeL PLACE 2010 WGCSE

| 105 | idh | 06/05/2010 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK-WEC-BTS |  |  |  |  |  |  |  |  |  |
| 1986 | 2009 |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.75 | 0.8 |  |  |  |  |  |  |
| 1 | 8 |  |  |  |  |  |  |  |  |
| 147.68 | 91 | 128 | 249 | 95 | 35 | 4 | 18 | 2 | 0 |
| 134.34 | 536 | 148 | 140 | 73 | 37 | 29 | 14 | 8 | 8 |
| 128.23 | 139 | 371 | 340 | 97 | 22 | 11 | 10 | 4 | 6 |
| 165.66 | 31 | 70 | 281 | 188 | 23 | 11 | 14 | 8 | 6 |
| 175.66 | 25 | 38 | 220 | 87 | 75 | 2 | 6 | 1 | 6 |
| 17168 | 22 | 27 | 63 | 79 | 62 | 41 | 9 | 0 | 1 |
| 196.6 | 152 | 44 | 72 | 24 | 40 | 20 | 17 | 3 | 5 |
| 189.19 | 21 | 70 | 60 | 24 | 13 | 25 | 13 | 11 | 2 |
| 205.87 | 34 | 32 | 98 | 30 | 10 | 2 | 9 | 13 | 8 |
| 187.15 | 50 | 46 | 45 | 48 | 12 | 4 | 5 | 6 | 1 |
| 184.37 | 33 | 106 | 30 | 17 | 25 | 5 | 1 | 3 | 7 |
| 184.74 | 53 | 122 | 197 | 24 | 6 | 12 | 7 | 1 | 1 |
| 185.49 | 81 | 125 | 125 | 85 | 9 | 6 | 7 | 4 | 0 |
| 187.89 | 38 | 44 | 182 | 53 | 30 | 3 | 2 | 6 | 4 |
| 180.37 | 48 | 63 | 125 | 179 | 38 | 22 | 1 | 2 | 0 |
| 177.98 | 32 | 64 | 51 | 111 | 97 | 25 | 13 | 0 | 3 |
| 179.74 | 138 | 102 | 87 | 23 | 23 | 40 | 5 | 2 | 0 |
| 182.24 | 29 | 137 | 60 | 50 | 5 | 18 | 27 | 7 | 0 |
| 163.99 | 11 | 33 | 59 | 23 | 10 | 3 | 1 | 10 | 0 |
| 186.6 | 30 | 75 | 91 | 70 | 13 | 3 | 3 | 5 | 2 |
| 184.74 | 55 | 102 | 103 | 30 | 31 | 3 | 4 | 0 | 5 |
| 18102 | 37 | 91 | 121 | 34 | 27 | 6 | 6 | 1 | 3 |
| 174.66 | 15 | 146 | 68 | 31 | 12 | 8 | 10 | 4 | 1 |
| 172.05 | 16 | 156 | 214 | 29 | 15 | 11 | 8 | 5 | 1 |
| UK-WECOT |  |  |  |  |  |  |  |  |  |
| 1988 | 2009 |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |
| 3 | 9 |  |  |  |  |  |  |  |  |
| 53.402 | 754.5 | 116.9 | 515 | 15.1 | 10 | 3.4 | 19 |  |  |
| 54.707 | 494 | 359.7 | 77 | 26.5 | 7 | 5.9 | 0.8 |  |  |
| 53.05 | 347.1 | 265.9 | 85.3 | 18.4 | 11.3 | 6 | 2.8 |  |  |
| 40.789 | 89.5 | 134.9 | 64.8 | 30.3 | 6.3 | 2.7 | 19 |  |  |
| 39.909 | 71.7 | 46.3 | 40.1 | 25.5 | 12.9 | 3.9 | 13 |  |  |
| 39.24 | 76.1 | 33.1 | 12 | 12.2 | 9.8 | 7.7 | 17 |  |  |
| 38.768 | 86.1 | 37.1 | 9.8 | 3.5 | 4.4 | 2.4 | 2.7 |  |  |
| 35.453 | 47.8 | 48.8 | 10.8 | 5.7 | 13 | 2.7 | 2.2 |  |  |
| 30.541 | 39.8 | 16.3 | 14.5 | 4 | 2 | 1 | 12 |  |  |
| 33.281 | 180.1 | 14.6 | 5.5 | 4.3 | 16 | 0.6 | 0.3 |  |  |
| 29.802 | 96.2 | 613 | 6.4 | 2.4 | 16 | 0.4 | 0.5 |  |  |
| 27.516 | 90.1 | 34.6 | 14.3 | 2.8 | 11 | 0.9 | 0.3 |  |  |
| 30.493 | 49.6 | 64.4 | 13.3 | 6.5 | 13 | 0.5 | 0.8 |  |  |
| 319 | 31.3 | 29.3 | 31.5 | 4.4 | 2.6 | 0.5 | 0.3 |  |  |
| 28.346 | 57.1 | 17.9 | 12.6 | 15.6 | 3.3 | 14 | 0.5 |  |  |
| 25.06 | 33.2 | 15.8 | 5.1 | 3.5 | 4.3 | 12 | 0.6 |  |  |
| 25.584 | 50.7 | 18.2 | 10.5 | 2.8 | 14 | 2.1 | 11 |  |  |
| 21129 | 24.1 | 17.6 | 5.7 | 2.6 | 0.8 | 0.8 | 0.8 |  |  |
| 21058 | 32.4 | 9.9 | 6.5 | 19 | 1 | 0.4 | 0.3 |  |  |
| 22.347 | 36.6 | 18.6 | 5.3 | 2.8 | 1 | 0.3 | 0.1 |  |  |
| 19.855 | 19.2 | 12.2 | 5.4 | 19 | 12 | 0.6 | 0.3 |  |  |
| 21398 | 43.8 | 8.6 | 3.5 | 18 | 0.7 | 0.5 | 0.1 |  |  |

Table 8.2.8 (Cont.) Plaice in VIle. Tuning fleet data available
(data in bold have been used fortuning)


Table 8.2.9 Plaice in VIIe. Diagnostics
Lowestoft VPA Version 3.1
10/05/2010 9:25
Extended Survivors Analysis
W.CHANNEL PLAICE 2010 WGCSE

CPUE data from file c: \vpa \PLE7ETU5.dat
Catch data for 30 years. 1980 to 2009. Ages 1 to 10 .

| Fleet | First <br> year | Last <br> year | First <br> age | Last <br> age | Alpha | Beta |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: |
| UK-WEC-BTS | 1986 | 2009 | 1 | 8 | 0.75 | 0.8 |
| UK WECOT | 1988 | 2009 | 3 | 9 | 0 | 1 |
| UK WECBT | 1989 | 2009 | 3 | 9 | 0 | 1 |
| UK WECOT historic | 1980 | 2009 | 2 | 9 | 0 | 1 |
| UK (E+W) FSP | 2003 | 2009 | 2 | 8 | 0.75 | 0.8 |

Time series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=7$

Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 5 years or the 4 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2.500$

Minimum standard error for population
estimates derived from each fleet $=.500$
Prior weighting not applied

Tuning converged after 27 iterations

| Regression weights | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 1 | 0.011 | 0.001 | 0.032 | 0.006 | 0.005 | 0.005 | 0.007 | 0 | 0.001 | 0.003 |
| 2 | 0.149 | 0.169 | 0.387 | 0.219 | 0.206 | 0.227 | 0.351 | 0.201 | 0.209 | 0.168 |
| 3 | 0.512 | 0.562 | 0.587 | 0.615 | 0.604 | 0.589 | 0.696 | 0.71 | 0.612 | 0.421 |
| 4 | 0.615 | 0.599 | 0.769 | 0.593 | 0.647 | 0.669 | 0.645 | 0.835 | 0.826 | 0.433 |
| 5 | 0.652 | 0.58 | 0.689 | 0.601 | 0.718 | 0.656 | 0.665 | 0.776 | 0.721 | 0.457 |
| 6 | 0.527 | 0.362 | 0.647 | 0.596 | 0.615 | 0.686 | 0.594 | 0.769 | 0.682 | 0.432 |
| 7 | 0.522 | 0.416 | 0.499 | 0.595 | 0.509 | 0.663 | 0.612 | 0.75 | 0.805 | 0.357 |
| 8 | 0.417 | 0.481 | 0.452 | 0.419 | 0.499 | 0.7 | 0.905 | 0.651 | 0.868 | 0.409 |
| 9 | 0.496 | 0.635 | 0.625 | 0.782 | 0.644 | 0.512 | 0.699 | 0.783 | 0.71 | 0.421 |

XSA population numbers (Thousands)

|  | AGE |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| YEAR |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |

$2000 \quad 4.55 \mathrm{E}+03 \quad 3.06 \mathrm{E}+03 \quad 3.53 \mathrm{E}+03 \quad 4.78 \mathrm{E}+03 \quad 1.10 \mathrm{E}+03 \quad 4.69 \mathrm{E}+02 \quad 9.90 \mathrm{E}+01 \quad 4.44 \mathrm{E}+01 \quad 5.86 \mathrm{E}+01$
$\begin{array}{lllllllll}2001 & 5.23 \mathrm{E}+03 & 3.99 \mathrm{E}+03 & 2.34 \mathrm{E}+03 & 1.87 \mathrm{E}+03 & 2.29 \mathrm{E}+03 & 5.08 \mathrm{E}+02 & 2.46 \mathrm{E}+02 & 5.21 \mathrm{E}+01\end{array} 2.60 \mathrm{E}+01$
$\begin{array}{llllllllll}2002 & 6.30 \mathrm{E}+03 & 4.63 \mathrm{E}+03 & 2.99 \mathrm{E}+03 & 1.18 \mathrm{E}+03 & 9.13 \mathrm{E}+02 & 1.14 \mathrm{E}+03 & 3.14 \mathrm{E}+02 & 1.44 \mathrm{E}+02 & 2.86 \mathrm{E}+01\end{array}$
$\begin{array}{lllllllll}2003 & 3.88 \mathrm{E}+03 & 5.41 \mathrm{E}+03 & 2.79 \mathrm{E}+03 & 1.48 \mathrm{E}+03 & 4.86 \mathrm{E}+02 & 4.07 \mathrm{E}+02 & 5.29 \mathrm{E}+02 & 1.69 \mathrm{E}+02\end{array} \quad 8.11 \mathrm{E}+01$
$2004 \quad 4.92 \mathrm{E}+03 \quad 3.42 \mathrm{E}+03 \quad 3.85 \mathrm{E}+03 \quad 1.34 \mathrm{E}+03 \quad 7.23 \mathrm{E}+02 \quad 2.36 \mathrm{E}+02 \quad 1.99 \mathrm{E}+02 \quad 2.59 \mathrm{E}+02 \quad 9.86 \mathrm{E}+01$
$2005 \quad 4.43 \mathrm{E}+03 \quad 4.34 \mathrm{E}+03 \quad 2.47 \mathrm{E}+03 \quad 1.87 \mathrm{E}+03 \quad 6.21 \mathrm{E}+02 \quad 3.13 \mathrm{E}+02 \quad 1.13 \mathrm{E}+02 \quad 1.06 \mathrm{E}+02 \quad 1.39 \mathrm{E}+02$
$2006 \quad 2.83 \mathrm{E}+03 \quad 3.91 \mathrm{E}+03 \quad 3.07 \mathrm{E}+03 \quad 1.21 \mathrm{E}+03 \quad 8.48 \mathrm{E}+02 \quad 2.86 \mathrm{E}+02 \quad 1.40 \mathrm{E}+02 \quad 5.18 \mathrm{E}+01 \quad 4.67 \mathrm{E}+01$
$2007 \quad 6.43 \mathrm{E}+03 \quad 2.50 \mathrm{E}+03 \quad 2.44 \mathrm{E}+03 \quad 1.36 \mathrm{E}+03 \quad 5.65 \mathrm{E}+02 \quad 3.87 \mathrm{E}+02 \quad 1.40 \mathrm{E}+02 \quad 6.72 \mathrm{E}+01 \quad 1.86 \mathrm{E}+01$
$2008 \quad 5.69 \mathrm{E}+03 \quad 5.70 \mathrm{E}+03 \quad 1.81 \mathrm{E}+031.06 \mathrm{E}+03 \quad 5.22 \mathrm{E}+02 \quad 2.31 \mathrm{E}+02 \quad 1.59 \mathrm{E}+02 \quad 5.87 \mathrm{E}+01 \quad 3.11 \mathrm{E}+01$
$2009 \quad 1.96 \mathrm{E}+03 \quad 5.04 \mathrm{E}+03 \quad 4.10 \mathrm{E}+03 \quad 8.70 \mathrm{E}+02 \quad 4.13 \mathrm{E}+02 \quad 2.25 \mathrm{E}+02 \quad 1.04 \mathrm{E}+02 \quad 6.31 \mathrm{E}+01 \quad 2.19 \mathrm{E}+01$

Table 8 2.s Plaice in Whe. Dingnostics (Cont.)

| Entinatei populationabuedance at htJan 20D |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0.008+00$ | 174E09 | 9.79E-09 | 2998409 | $5.01 \mathrm{t}+02$ | 292E+02 | $1908+02$ | 6422+01 | 9.7E+01 |  |
| Taperweriteig geometiric mean offile VPApopuhtiona: |  |  |  |  |  |  |  |  |  |  |
|  | 5.73E408 | 5925409 | 99E-09 | 190E409 | 860E402 | 4.15E+02 | $2178+02$ | 18E+02 | 600E401 |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | 0.4954 | 0.4989 | 05007 | 05706 | 05578 | 05821 | 0.616: | 06582 | 0.769 |  |
| Logcatclubinymenitale. |  |  |  |  |  |  |  |  |  |  |
| Flet:LE-WEC-BTS |  |  |  |  |  |  |  |  |  |  |
| $\begin{array}{rr}\text { Afe } \\ & \\ \\ & 2 \\ \\ \\ \\ 4 \\ & 5 \\ & 6 \\ & 7 \\ & \\ & \\ & \end{array}$ | 1880 | 181 | 182 | BES | E84 | 1385 | 1886 | 1887 | 198 | 1989 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | -09 | 179 | 0.8 | -0.1 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | 0.9 | -0.92 | 0.94 | -0.78 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | 0.61 | 0.07 | 0.26 | 001 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | 042 | 0.28 | 0.96 | 0.5 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | 0.17 | 0.51 | -0.07 | .0.87 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | -0.57 | 0.86 | 0.4 | -0.09 |
|  | 99.99 | 9999 | 9999 | 9999 | 9999 | 9999 | 085 | 12 | 0.02 | 0.97 |
|  | 9999 | 9999 | 9999 | 9999 | 9999 | 9999 | -087 | 0.68 | 0.59 | 0.8 |
|  | 9 Mo data forte fleet at mer |  |  |  |  |  |  |  |  |  |
| $\mathbf{A F E}^{\text {e }}$ | 199 | 1991 | B92 | B99 | B94 | 1995 | 1996 | 1997 | 1988 | 199 |
|  | -0.44 | -0.67 | 099 | -0.E | $0 \times 1$ | -0.09 | -062 | -0.08 | 0.57 | 0.29 |
|  | -0.54 | -0.86 | -0.69 | -0.28 | -0.96 | 0.06 | -0.09 | $0 \mathbf{O}$ | -0.94 | -0.59 |
|  | 0.04 | -0.26 | -0.2 | -0.51 | -0.0.2 | -0.4 | -0.69 | 0.95 | -0.5 | -0.4 |
|  | -0.45 | -0.1 | -052 | -0.47 | -0.41 | -0.09 | -099 | -0.05 | 0.94 | -0.26 |
|  | 0.02 | 0.11 | -002 | -0.2 | -0.58 | -0.0.4 | 022 | -0.44 | -0.2 | 0.18 |
|  | -192 | 0.29 | -0.24 | 0.49 | -127 | -0.5 | -029 | 0.49 | 0.47 | -0.95 |
|  | -0.97 | -0.5 | -0.48 | -0.91 | -0.44 | 0.06 | -154 | 0.56 | 0.6 | -0.25 |
|  | -167 | 9999 | -0.82 | -0.28 | 0.1 | -0.6 | 0.97 | -0.86 | 0.69 | 0.59 |
|  | 9 No data for tie fleet at tim age |  |  |  |  |  |  |  |  |  |
| $A_{5}{ }^{\text {e }}$ | 2000 | 2001 | 2002 | 2009 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 0.24 | -09 | 099 | -0.2 | -122 | -0.24 | 082 | -0.88 | -138 | 0.02 |
|  | 0.22 | 0 | 0.47 | 0.47 | -0.4 | 0.07 | 059 | 0.85 | 0.52 | 0.69 |
|  | 0.17 | -0.26 | 004 | -0.26 | -0.5 | 0.24 | 024 | 0.66 | 0.34 | 0.54 |
|  | 0.29 | 0.75 | -0.24 | 0.6 | -097 | 09 | -0.18 | 0.05 | 0.39 | 0.08 |
|  | 0.96 | 052 | 0.0 | -0.9 | -0.41 | -0.17 | 04 | 0.78 | 0.04 | 091 |
|  | 08 | 0.74 | 0.61 | 0.79 | -0.94 | -0.69 | -067 | -0.2 | 0.65 | 0.82 |
|  | -105 | 054 | -0.61 | 0.62 | -166 | -0.01 | 004 | 0.57 | 109 | 0.91 |
|  | 0.97 | 9999 | -0.72 | 0.77 | 0.97 | 0.6 | 9999 | -0.56 | 16 | 0.97 |


orepenient of yearchn: inemath and com tant writine

| $A_{5}$ | 1 | 2 | g | 4 | 5 | 6 | 7 | 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Meamiog 4 | -98816 | -9.0999 | \$.287 | EPES | \$ 3241 | -557 | - $\$ 2469$ | -2469 |
| SE(OE $)^{\text {] }}$ | 0.711 | 05B6 | 0.9609 | 09978 | 0.997 | 0.72 | 0.7421 | 0729 |

Regrenionentatiofici :



| 1 | 0.87 | 0596 | 9.71 | 0.42 | 24 | 0.69 | -985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 104 | -0.169 | 911 | 0.45 | 24 | 055 | -909 |
| 9 | 0.91 | 0.669 | 322 | 0.72 | 24 | 0.9 | - 22 |
| 4 | 0.89 | 1049 | 8.8 | 0.79 | 24 | 0.9 | $\underline{2}$ |
| 5 | 0.82 | 1602 | 805 | 0.78 | 24 | 0.91 | - 82 |
| 6 | 0.88 | 0.498 | 8.27 | 0.45 | 24 | 0.65 | - 25 |
| 7 | 108 | -0.17 | 893 | 0.8 | 24 | 0.78 | -25 |
| E | 15 | -1747 | 1068 | 022 | 21 | 121 | - 2.21 |

## Table 8.2.9 Plaice in VIle. Diagnostics (Cont.

| Fleet:UK WECOT |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
|  | 1 No data for this fleetatthis age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.57 | 0.37 |
|  | 4 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.09 | 0.52 |
|  | 5 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.38 | 0.58 |
|  | 6 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.08 | 0.56 |
|  | 7 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.13 | -0.02 |
|  | 8 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.44 | 0.12 |
|  | 9 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.14 | -0.29 |
| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | 1 No data for this fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleet at this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.32 | 0.15 | -0.01 | -0.07 | -0.11 | 0.2 | 0.09 | 0.57 | 0.08 | -0.49 |
|  | 4 | 0.45 | 0.44 | 0.3 | 0.03 | 0.03 | 0.18 | 0.03 | -0.25 | 0.43 | -0.17 |
|  | 5 | 0.05 | 0.29 | 0.27 | -0.02 | -0.23 | -0.08 | 0.14 | -0.15 | 0 | 0.05 |
|  | 6 | 0.18 | 0.13 | 0.27 | -0.01 | -0.35 | 0.2 | -0.03 | -0.21 | 0.08 | 0.16 |
|  | 7 | 0.63 | 0.15 | 0.05 | 0.17 | -0.26 | -0.44 | 0.13 | -0.1 | -0.28 | 0.23 |
|  | 8 | 0.49 | 0.06 | 0.24 | 0.13 | -0.68 | -0.11 | 0.17 | -0.49 | -0.64 | -0.19 |
|  | 9 | 0.05 | 0.23 | -0.14 | 0 | -0.32 | -0.16 | -0.15 | -0.23 | 0.22 | -0.19 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 No data for this fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 | -0.31 | -0.39 | 0.1 | -0.24 | -0.16 | -0.28 | -0.15 | 0.15 | -0.12 | -0.28 |
|  | 4 | -0.33 | -0.23 | -0.07 | -0.37 | -0.13 | -0.29 | -0.45 | 0.1 | 0.03 | -0.36 |
|  | 5 | -0.23 | -0.18 | -0.01 | -0.2 | 0.16 | -0.14 | -0.31 | -0.12 | 0.07 | -0.32 |
|  | 6 | 0.05 | -0.54 | 0.16 | -0.2 | 0.11 | -0.02 | -0.28 | -0.18 | 0.03 | -0.19 |
|  | 7 | 0.17 | -0.13 | 0.02 | -0.08 | -0.28 | -0.02 | -0.02 | -0.02 | 0.17 | -0.21 |
|  | 8 | -0.03 | -0.2 | -0.08 | -0.29 | -0.14 | 0.07 | 0.18 | -0.54 | 0.5 | -0.03 |
|  | 9 | 0.2 | 0.05 | 0.58 | -0.09 | 0.24 | -0.29 | -0.09 | -0.29 | 0.38 | -0.57 |

Mean log catc hability and standard erro r of ages with catc hability
independent of yearclass strength and constantw.r.t. time

| Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean $\log q$ | -7.0677 | -7.0535 | -7.2397 | -7.4413 | -7.6209 | -7.6209 | -7.6209 |
| S.E $\log \mathrm{G})$ | 0.292 | 0.2942 | 0.2324 | 0.2395 | 0.2278 | 0.3406 | 0.2731 |

Regression statistics:

Ages with q independent of yearclass strength and constant w.r.t. time.

| Age |  |  | $t$-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 0.82 | 1.905 | 7.27 | 0.85 | 22 | 0.23 | -7.07 |
|  | 4 | 0.79 | 2.704 | 7.15 | 0.9 | 22 | 0.2 | -7.05 |
|  | 5 | 0.84 | 2.461 | 7.16 | 0.92 | 22 | 0.18 | -7.24 |
|  | 6 | 0.89 | 1.501 | 7.28 | 0.9 | 22 | 0.21 | -7.44 |
|  | 7 | 0.95 | 0.597 | 7.52 | 0.89 | 22 | 0.22 | -7.62 |
|  | 8 | 0.98 | 0.185 | 7.61 | 0.8 | 22 | 0.34 | -7.67 |
|  | 9 | 1.04 | -0.506 | 7.81 | 0.88 | 22 | 0.29 | -7.67 |

Fleet:UK WECBT

Age

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | No data forthis fleetat this age |  |  |  |  |  |  |  |  |  |
| 2 | No data forthis fleetat this age |  |  |  |  |  |  |  |  |  |
| 3 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.34 |
| 4 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.17 |
| 5 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.08 |
| 6 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.04 |
| 7 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.01 |
| 8 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | -0.03 |
| 9 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 0.01 |

Table 8.2.9 Plaice in VIle. Diagnostics (Cont.)

| Age |  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 No data forthis fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 | No data forthis | is age |  |  |  |  |  |  |  |  |
|  | 3 | 0.15 | 0.2 | 0.52 | 0.5 | 0.34 | 0.31 | -0.17 | 0.09 | -0.42 | -0.34 |
|  | 4 | 0.17 | 0.5 | 0.72 | 0.51 | 0.44 | 0.31 | -0.13 | -0.54 | -0.08 | -0.2 |
|  | 5 | -0.21 | 0.32 | 0.59 | 0.41 | 0.22 | 0.15 | 0 | -0.21 | -0.25 | 0.01 |
|  | 6 | -0.15 | -0.04 | 0.36 | 0.35 | 0.09 | 0.24 | -0.16 | -0.12 | 0.03 | 0.37 |
|  | 7 | 0.31 | -0.07 | 0.04 | 0.38 | 0.04 | -0.05 | -0.07 | 0.04 | -0.02 | 0.24 |
|  | 8 | -0.03 | -0.47 | 0.18 | 0.28 | -0.2 | 0.15 | 0.13 | -0.14 | -0.11 | 0.12 |
|  | 9 | -0.29 | -0.28 | -0.07 | 0.13 | 0.13 | 0.28 | 0.05 | 0.24 | 0.53 | 0.52 |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 1 No data forthis fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 No data for this fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 3 | -0.25 | -0.45 | 0.03 | -0.03 | -0.02 | 0.03 | 0.09 | -0.01 | -0.23 | -0.01 |
|  | 4 | -0.36 | -0.24 | -0.22 | -0.2 | 0.07 | -0.07 | -0.24 | -0.05 | -0.09 | -0.13 |
|  | 5 | -0.42 | -0.12 | -0.1 | -0.09 | 0.14 | 0.06 | -0.11 | -0.1 | -0.1 | -0.13 |
|  | 6 | -0.3 | -0.53 | -0.01 | -0.02 | 0.05 | 0.16 | -0.12 | -0.07 | -0.09 | -0.09 |
|  | 7 | -0.29 | -0.28 | -0.17 | 0.13 | 0.01 | 0.17 | 0.01 | 0.09 | -0.09 | -0.4 |
|  | 8 | -0.25 | -0.25 | -0.3 | -0.28 | -0.01 | 0.17 | 0.23 | 0.05 | 0.34 | 0.02 |
|  | 9 | -0.26 | -0.1 | -0.06 | 0.46 | 0.29 | -0.01 | -0.05 | 0.27 | 0.1 | 0.13 |

Mean log catc hability and standard error of ages with catchability
independent of yearclass strength and constant w.r.t. time

| Age |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean $\operatorname{Logq}$ | -6.4303 | -6.278 | -6.279 | -6.3228 | -6.3802 | -6.3802 | -6.3802 |
| S.E(Log $)$ | 0.2793 | 0.3234 | 0.2381 | 0.2186 | 0.1911 | 0.2184 | 0.2619 |

Regression statistics :

Ages with $q$ independent of yearclass strength and constant w.r.t. time.

| Age | Slope | t-value |  | Intercept | RSquare | No Pts | Regs.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |

Fleet: UK WECOT historic
Age

|  | 1980 | 1981 |
| :---: | :---: | ---: |
| 1 | No data forthis fleetat this age |  |
| 2 | -0.16 | 0.08 |
| 3 | -0.25 | 0.26 |
| 4 | -0.37 | -0.02 |
| 5 | -0.34 | -0.02 |
| 6 | 0.38 | -0.11 |
| 7 | -0.41 | 0.15 |
| 8 | -0.39 | 0.16 |
| 9 | 0 | 0.23 |

1982

-0.06
0.02
0.22
0.06
0.29
-0.02
0.54
0.11
1983

0.25
0.08
0.35
0.5
0
0.19
-0.05
-0.24

| 1984 | 1985 |
| :---: | :---: |
|  |  |
| 0.54 | -0.3 |
| -0.06 | -0.18 |
| 0.11 | -0.06 |
| 0.08 | -0.5 |
| -0.09 | 0.11 |
| 0.23 | 0.11 |
| 0.34 | -0.57 |
| 0.17 | -0.41 |

1986

0.09
0.12
-0.4
-0.18
-0.53
0.02
-0.04
0.3
1987
-0.43

## Table 8.2.9 Plaice in VIle. Diagnostics (Cont.)

| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 No data for this fleetat this age |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 3 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 4 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 5 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 6 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 7 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 8 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 9 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |

Mean $\log$ catc hability and standard error of ages with catc hability
independent of year class strength and constantw.r.t.t time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -7.2676 | -5.9594 | -5.8043 | -5.9658 | -6.0674 | -5.9803 | -5.9803 | -5.9803 |
| S.E(Log q) | 0.3113 | 0.1647 | 0.2703 | 0.3443 | 0.2807 | 0.226 | 0.3873 | 0.2496 |

Regression statistics:

Ages with q independent of yearclass strength and constant w.r.t. time.

| Age |  |  | $t$-value | Intercept | RSquare | No Pts |  | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 1.43 | -1.13 | 6.58 | 0.54 |  | 8 | 0.44 | -7.27 |
|  | 3 | 0.83 | 1.321 | 6.37 | 0.91 |  | 8 | 0.13 | -5.96 |
|  | 4 | 0.79 | 1.59 | 6.18 | 0.91 |  | 8 | 0.19 | -5.8 |
|  | 5 | 0.73 | 1.511 | 6.18 | 0.84 |  | 8 | 0.23 | -5.97 |
|  | 6 | 1.33 | - 1.471 | 6.11 | 0.77 |  | 8 | 0.34 | -6.07 |
|  | 7 | 1.12 | -0.769 | 6.07 | 0.87 |  | 8 | 0.26 | -5.98 |
|  | 8 | 1.48 | - 1.603 | 6.71 | 0.65 |  | 8 | 0.51 | -6.03 |
|  | 9 | 0.81 | 2.871 | 5.54 | 0.97 |  | 8 | 0.14 | -5.94 |

Fleet:UK $(E+W)$ FSP

Age

|  | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: |
| 1 | No data forthis fleetat this age |  |  |  |
| 2 | 99.99 | 99.99 | 99.99 | -0.3 |
| 3 | 99.99 | 99.99 | 99.99 | -0.0 |
| 4 | 99.99 | 99.99 | 99.99 | -0.2 |
| 5 | 99.99 | 99.99 | 99.99 | -0.0 |
| 6 | 99.99 | 99.99 | 99.99 | 0.0 |
| 7 | 99.99 | 99.99 | 99.99 | 0.2 |
| 8 | 99.99 | 99.99 | 99.99 | 0.3 |
| 9 | No data forthis fleetatthis age |  |  |  |


| 2004 | 2005 | 2006 |
| ---: | ---: | ---: |
|  |  |  |
| 0.13 | -0.07 | 0.64 |
| 0.19 | 0.17 | 0.07 |
| 0.21 | -0.04 | 0.08 |
| 0.37 | 0.06 | -0.12 |
| -0.19 | 0.27 | 0.09 |
| 0.26 | -0.39 | 0.35 |
| 0.2 | 0.17 | 0.07 |


| 2008 | 2009 |
| ---: | ---: |
|  |  |
| 99.99 | -0.06 |
| 99.99 | -0.44 |
| 99.99 | 0.11 |
| 99.99 | -0.25 |
| 99.99 | -0.05 |
| 99.99 | 0.1 |
| 99.99 | -0.51 |

Mean $\log$ catchability and standard error of ages with catchability
independent of yearclass strength and constantw.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean $\log q$ | -9.2562 | -8.448 | -8.3312 | -8.3455 | -8.5181 | -8.3363 | -8.3363 |
| S.E(Log $q)$ | 0.3582 | 0.2317 | 0.1687 | 0.2114 | 0.1678 | 0.3716 | 0.3511 |

Regression statistics :

Ages with q independent of yearc lass strength and constant w.r.t. time.

| Age |  |  | $t$-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 0.96 | 0.068 | 9.22 | 0.4 | 6 | 0.38 | -9.26 |
|  | 3 | 2.18 | -1.21 | 8.95 | 0.21 | 6 | 0.48 | -8.45 |
|  | 4 | 1.43 | -1.008 | 8.83 | 0.57 | 6 | 0.24 | -8.33 |
|  | 5 | 0.71 | 1.138 | 7.79 | 0.8 | 6 | 0.15 | -8.35 |
|  | 6 | 0.89 | 0.348 | 8.22 | 0.73 | 6 | 0.17 | -8.52 |
|  | 7 | 0.81 | 0.821 | 7.73 | 0.82 | 6 | 0.31 | -8.34 |
|  | 8 | 0.73 | 1.787 | 7.34 | 0.92 | 6 | 0.21 | -8.35 |

## Table 8.2.9 Plaice in VIle. Diagnostics (Cont.)

Terminal yearsurvivor and F summaries:

Age 1 Catchability constantw.r.t. time and dependenton age

Yearclass $=2008$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| UK-WEC-BTS | 1770 | 0.726 | 0 | 0 | 1 | 0.922 | 0.003 |
| UK WECOT | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK WECBT | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK WECOT historic | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (E+W) FSP | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 1375 | 2.5 |  |  |  | 0.078 | 0.004 |

Weighted prediction :


Age 2 Catchability constantw.r.t. time and dependenton age
Yearclass $=2007$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| UK-WEC-BTS | 4004 | 0.428 | 0.865 | 2.02 | 2 | 0.565 | 0.16 |
| UK WECOT | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK WECBT | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK WECOT historic | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (E+W) FSP | 3552 | 0.5 | 0 | 0 | 1 | 0.415 | 0.178 |
| F shrinkage mean | 2558 | 2.5 |  |  |  | 0.02 | 0.24 |

Weighted prediction :


Age 3 Catchability constantw.r.t. time and dependenton age

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sunivors | s.e | S.e | Ratio |  | Weights | F |
| UK-WEC-BTS | 3432 | 0.327 | 0.249 | 0.76 | 3 | 0.408 | 0.311 |
| UK WECOT | 1810 | 0.5 | 0 | 0 | 1 | 0.194 | 0.525 |
| UK WECBT | 2357 | 0.5 | 0 | 0 | 1 | 0.194 | 0.426 |
| UK WECOT historic | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| UK (E+W) FSP | 1541 | 0.5 | 0 | 0 | 1 | 0.194 | 0.594 |
| F shrinkage mean | 1379 | 2.5 |  |  |  | 0.012 | 0.646 |

Weighted prediction:


## Table 8.2.9 Plaice in VIle. Diagnostics (Cont.)

Age 4 Catchability constantw.r.t. time and dependenton age

Yearclass $=2005$

| Fleet | Estimated |  | Int | Ext | Var | $N$ | Scaled <br> Weights |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
|  | Survivors |  | S.e | S.e | Ratio | F |  |

Weighted prediction :

| Survivors |  | Int | Ext | N |  | Var | F |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| atend of year | S.e | S.e |  |  | Ratio |  |  |  |
|  | 501 |  | 0.17 |  | 0.11 | 11 | 0.613 | 0.433 |

Age 5 Catchability constantw.r.t. time and dependenton age
Yearclass $=2004$

| Fleet | Estimated |  | Int | Ext | Var | N | Scaled <br> Weights |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | :---: |
| Sunvivors |  | S.e | s.e | Ratio | F |  |  |

Weighted prediction :

| Sunivors |  | Int | Ext | N |  | Var | F |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :--- |
| atend ofyear |  | S.e | S.e |  |  | Ratio |  |  |
|  | 232 |  | 0.17 |  | 0.08 | 15 | 0.454 | 0.457 |

Age 6 Catchability constantw.r.t. time and dependenton age

Yearclass $=2003$


## Table 8.2.9 Plaice in VIle. Diagnostics (Cont.)

Age 7 Catchability constantw.r.t. time and dependent on age

Yearclass $=2002$

| Fleet | Estimated |  | Int | Ext | Var | N | Scaled |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | :---: |
|  | Survivors |  | s.e | S.e | Ratio |  | Weights |

Weighted prediction :


Age 8 Catchability constantw.r.t. time and age (fixed at the value forage) 7
Yearclass $=2001$

| Fleet | Estimated |  | Int | Ext | Var | $N$ | Scaled <br> Weights |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
|  | Survivors |  | S.estimated |  |  |  |  |

Weighted prediction :


Age 9 Catchability constantw.r.t. time and age (fixed at the value forage) 7

Yearclass $=2000$

| Fleet | Estimated |  | Int | Ext | Var | N | Scaled <br> Weights |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
|  | Survivors |  | s.e | s.e | Ratio |  | Fstimated |

Weighted prediction :

| Survivors at end of year |  | Int |  | Ext |  | $N$ |  | Var |  | F |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | s.e |  | s.e |  |  |  | Ratio |  |  |  |
|  | 13 |  | 0.18 |  | 0.08 |  | 28 |  | 0.471 |  | 0.421 |

Table 8.2.10 Plaice in VIIe. Fishing mortality-at-age.
Run title : W.CHANNEL PLAICE 2010 WGCSE

At 10/05/2010 9:27

Terminal Fs derived using XSA (With F shrinkage)

| YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.0024 | 0.0120 | 0.0098 | 0.0005 | 0.0097 | 0.0004 | 0.0006 | 0.0055 | 0.0012 | 0.0024 |
| 2 | 0.1241 | 0.1086 | 0.1091 | 0.1303 | 0.1832 | 0.1091 | 0.1505 | 0.0880 | 0.1772 | 0.0413 |
| 3 | 0.4328 | 0.5360 | 0.4799 | 0.4684 | 0.4627 | 0.4702 | 0.6026 | 0.5297 | 0.5572 | 0.4079 |
| 4 | 0.4918 | 0.5995 | 0.6947 | 0.8086 | 0.7539 | 0.6887 | 0.5208 | 0.7299 | 0.5053 | 0.6721 |
| 5 | 0.4276 | 0.4155 | 0.5302 | 0.6390 | 0.5758 | 0.3338 | 0.4936 | 0.7451 | 0.5098 | 0.7740 |
| 6 | 0.7298 | 0.3145 | 0.5650 | 0.3734 | 0.4073 | 0.5465 | 0.4017 | 0.4206 | 0.3496 | 0.5865 |
| 7 | 0.3457 | 0.5119 | 0.4076 | 0.5057 | 0.6062 | 0.5258 | 0.5376 | 0.5668 | 0.3477 | 0.4598 |
| 8 | 0.3902 | 0.4270 | 0.9835 | 0.4169 | 0.7718 | 0.3413 | 0.4648 | 0.3891 | 0.4824 | 0.4999 |
| 9 | 0.4623 | 0.5728 | 0.4353 | 0.2918 | 0.4870 | 1.1677 | 0.6151 | 0.5871 | 0.2992 | 0.8048 |
| tgp | 0.4623 | 0.5728 | 0.4353 | 0.2918 | 0.4870 | 1.1677 | 0.6151 | 0.5871 | 0.2992 | 0.8048 |
| FBAR 3.6 | 0.5205 | 0.4664 | 0.5675 | 0.5724 | 0.5499 | 0.5098 | 0.5047 | 0.6063 | 0.4805 | 0.6101 |


| YEAR | $1990$ | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.0127 | 0.0080 | 0.0154 | 0.0134 | 0.0298 | 0.0008 | 0.0022 | 0.0007 | 0.0014 | 0.0059 |
| 2 | 0.1114 | 0.1916 | 0.2067 | 0.1787 | 0.1894 | 0.1876 | 0.1903 | 0.1765 | 0.0720 | 0.1664 |
| 3 | 0.6162 | 0.6150 | 0.6791 | 0.6130 | 0.5912 | 0.6354 | 0.5585 | 0.6885 | 0.4968 | 0.3977 |
| 4 | 0.7526 | 0.7935 | 0.7734 | 0.6898 | 0.8265 | 0.7448 | 0.6790 | 0.7540 | 0.7540 | 0.7151 |
| 5 | 0.5767 | 0.5914 | 0.6149 | 0.6170 | 0.5812 | 0.5879 | 0.6835 | 0.6979 | 0.5355 | 0.6050 |
| 6 | 0.5534 | 0.4780 | 0.5498 | 0.5084 | 0.5112 | 0.5390 | 0.5531 | 0.5753 | 0.5216 | 0.6363 |
| 7 | 0.5844 | 0.3924 | 0.4247 | 0.5119 | 0.3973 | 0.5041 | 0.5195 | 0.7458 | 0.4269 | 0.5933 |
| 8 | 0.5648 | 0.3595 | 0.4620 | 0.4859 | 0.3348 | 0.4876 | 0.7380 | 0.5770 | 0.4475 | 0.4977 |
| 9 | 0.5597 | 0.4465 | 0.4318 | 0.4514 | 0.4411 | 0.4696 | 0.5203 | 0.7942 | 0.7747 | 0.6370 |
| tgp | 0.5597 | 0.4465 | 0.4318 | 0.4514 | 0.4411 | 0.4696 | 0.5203 | 0.7942 | 0.7747 | 0.6370 |
| FBAR 3.6 | 0.6247 | 0.6195 | 0.6543 | 0.6071 | 0.6276 | 0.6268 | 0.6185 | 0.6789 | 0.5770 | 0.5885 |



Table 8.2.11 Plaice in VIIe. Stock numbers-at-age.

| Run tite: W.Channel place 2010 w |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| At100552010 9:27 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Teminal Fs deived using XSA (With F shinkage) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Table 10 Stocknumberatage (statofyear) Numbers.00**3 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| year | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |
| age |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 8432 | 3637 | 7814 | 6942 | 8504 | 8791 | 17880 | 14321 | 10433 | 4450 |  |  |  |
| 2 | 7407 | 7461 | 3187 | 6863 | 6154 | 7470 | 7794 | 15849 | 12632 | 9242 |  |  |  |
| 3 | 2419 | 5803 | 5936 | 2534 | 5343 | 4544 | 5941 | 5947 | 12872 | 9384 |  |  |  |
| 4 | 690 | 1392 | 3011 | 3258 | 1407 | 2983 | 2519 | 2884 | 3106 | 6540 |  |  |  |
| 5 | 700 | 374 | 678 | 1333 | 128 | 587 | 1329 | 1327 | 1233 | 1662 |  |  |  |
|  | 128 | 405 | 219 | 354 | 624 | 642 | 373 | 719 | 559 | 657 |  |  |  |
| 7 | 229 | 55 | 262 | 110 | 216 | 368 | 330 | 221 | 419 | 349 |  |  |  |
| 8 | 76 | 144 | 29 | 155 | 59 | 104 | 193 | 171 | 111 | 262 |  |  |  |
| 9 | 38 | 46 | 83 | 10 | 90 | 24 | 66 | 108 | 103 | 61 |  |  |  |
| tgp | 394 | 231 | 364 | 417 | 138 | 79 | 141 | 123 | 217 | 192 |  |  |  |
| TOTAL | 20513 | 19546 | 21584 | 21975 | 23823 | 25594 | 36565 | 41671 | 41685 | 32800 |  |  |  |
| Table 10 Stocknumberatage (statofyear) |  | Numbers*10*3 |  |  |  |  |  |  |  |  |  |  |  |
| year | 1990 | 199 | 1992 | 1993 | 199 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 4803 | 5434 | 6269 | 2874 | 3033 | 8022 | 7139 | 10973 | 5302 | 3470 |  |  |  |
| 2 | 3938 | 4206 | 4781 | 5475 | 2515 | 2611 | 710 | 6317 | 9725 | 4696 |  |  |  |
| 3 | 7866 | 3124 | 3080 | 3448 | 4062 | 1846 | 1920 | 5213 | 4696 | 8027 |  |  |  |
| 4 | 5535 | 3767 | 1498 | 1885 | 1657 | 1994 | 867 | 974 | 2323 | 2534 |  |  |  |
|  | 2962 | 2313 | 1511 | 613 | 616 | 643 | 840 | 390 | 406 | 969 |  |  |  |
| 6 | 680 | 1476 | 1136 | 725 | 293 | 306 | 317 | 376 | 172 | 211 |  |  |  |
| 7 | 324 | 347 | 811 | 581 | 387 | 156 | 158 | 162 | 188 | 91 |  |  |  |
| 8 | 196 | 160 | 208 | 471 | 309 | 230 | 84 | 83 | 68 | 109 |  |  |  |
| 9 | 141 | 99 | 99 | 116 | 257 | 196 | 126 | 35 | 42 | 39 |  |  |  |
| tgp | 179 | 151 | 151 | 249 | 192 | 252 | 279 | 194 | 122 | 102 |  |  |  |
| TOTAL | 26623 | 21077 | 19544 | 15938 | 13321 | 1625 | 18838 | 24717 | 23044 | 20248 |  |  |  |
| Table 10 Stocknumberatage (statofyear) |  | Numbers*00*.3 |  |  |  |  |  |  |  |  |  |  |  |
| year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 4553 | 5230 | 6297 | 3879 | 4916 | 4428 | 2833 | 6434 | 5686 | 1962 | 0 | 5981 | 6682 |
| 2 | 3060 | 3995 | 4633 | 5408 | 3419 | 4340 | 3906 | 2495 | 5704 | 5039 | 1736 | 5316 | 5953 |
| 3 | 3526 | 2338 | 2992 | 2791 | 3851 | 2467 | 3067 | 2439 | 1810 | 4104 | 3776 | 4013 | 4553 |
| 4 | 4783 | 1874 | 1182 | 1476 | 1338 | 1868 | 1214 | 1356 | 1063 | 870 | 2389 | 2000 | 2336 |
| 5 | 1099 | 2294 | 913 | 486 | 723 | 621 | 848 | 565 | 522 | 413 | 501 | 899 | 1047 |
| 6 | 469 | 508 | 1140 | 407 | 236 | 313 | 286 | 387 | 231 | 225 | 232 | 427 | 504 |
| 7 | 99 | 246 | 314 | 529 | 199 | 113 | 140 | 140 | 159 | 104 | 130 | 225 | 269 |
| 8 | 44 | 52 | 144 | 169 | 259 | 106 | 52 | 67 | 59 | 63 | 64 | 120 | 147 |
| 9 | 59 | 26 | 29 | 81 | 99 | 139 | 47 | 19 | 31 | 22 | 37 | 64 | 81 |
| tgp | 141 | 83 | 95 | 76 | 85 | 115 | 94 | 94 | 59 | 49 | 41 |  |  |
| TOTAL | 17834 | 16646 | 1738 | 15302 | 1512 | 14511 | 1288 | 13998 | 15325 | 12850 | 8906 |  |  |

## Table 8.2.12 Plaice in VIle. Summary

Run title : W.CHANNEL PLAICE 2010 WGCSE

| At 10/05/2010 9:27 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Table 16 | 6 Summary | (without SOP correction) |  |  |  |  |
|  | Terminal Fs derived using XSA (With F shrinkage) |  |  |  |  |  |
|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 3-6 |
|  | Age 1 |  |  |  |  |  |
| 1980 | 8432 | 5051 | 2410 | 1215 | 0.5043 | 0.5205 |
| 1981 | 3637 | 6256 | 3282 | 1746 | 0.532 | 0.4664 |
| 1982 | 7814 | 5901 | 3470 | 1938 | 0.5586 | 0.5675 |
| 1983 | 6942 | 6234 | 3665 | 1754 | 0.4787 | 0.5724 |
| 1984 | 8504 | 6386 | 3487 | 1813 | 0.5198 | 0.5499 |
| 1985 | 8791 | 6680 | 3564 | 1751 | 0.4912 | 0.5098 |
| 1986 | 17880 | 7582 | 3753 | 2161 | 0.5759 | 0.5047 |
| 1987 | 14321 | 7095 | 3626 | 2388 | 0.6586 | 0.6063 |
| 1988 | 10433 | 9821 | 5166 | 2994 | 0.5796 | 0.4805 |
| 1989 | 4450 | 9002 | 5487 | 2808 | 0.5119 | 0.6101 |
| 1990 | 4803 | 8598 | 5297 | 3058 | 0.5773 | 0.6247 |
| 1991 | 5434 | 6653 | 4313 | 2250 | 0.5216 | 0.6195 |
| 1992 | 6269 | 6568 | 3595 | 1950 | 0.5424 | 0.6543 |
| 1993 | 2874 | 5158 | 3069 | 1691 | 0.5511 | 0.6071 |
| 1994 | 3033 | 4455 | 2721 | 1471 | 0.5405 | 0.6276 |
| 1995 | 8022 | 4872 | 2420 | 1295 | 0.535 | 0.6268 |
| 1996 | 7139 | 4918 | 2376 | 1321 | 0.5561 | 0.6185 |
| 1997 | 10973 | 6420 | 2505 | 1654 | 0.6604 | 0.6789 |
| 1998 | 5302 | 5881 | 2669 | 1430 | 0.5358 | 0.577 |
| 1999 | 3470 | 4964 | 2960 | 1616 | 0.546 | 0.5885 |
| 2000 | 4553 | 4800 | 3293 | 1678 | 0.5096 | 0.5765 |
| 2001 | 5230 | 4456 | 2722 | 1379 | 0.5066 | 0.5256 |
| 2002 | 6297 | 4923 | 2510 | 1608 | 0.6406 | 0.6729 |
| 2003 | 3879 | 4255 | 2507 | 1478 | 0.5895 | 0.6015 |
| 2004 | 4916 | 4873 | 2274 | 1402 | 0.6166 | 0.646 |
| 2005 | 4428 | 4521 | 2257 | 1370 | 0.6071 | 0.6502 |
| 2006 | 2833 | 3822 | 2059 | 1466 | 0.7118 | 0.6498 |
| 2007 | 6434 | 3574 | 1717 | 1184 | 0.6896 | 0.7727 |
| 2008 | 5686 | 4009 | 1653 | 1144 | 0.6919 | 0.7104 |
| 2009 | 1962 | 3161 | 1833 | 1043 | 0.5691 | 0.4356 |
| Arith. |  |  |  |  |  |  |
| Mean | 6491 | 5696 | 3089 | 1735 | 0.5703 | 0.5951 |
| Units | (Thousands | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 8.2.13 VIle plaice : Catch forecast input data

MFDP version 1a
Run: ple7e2010
Time and date: 14:07 16/05/2010
Fbar age range: 3-6

2010

| Age | $\boldsymbol{N}$ | $\boldsymbol{M}$ | Mat | $\boldsymbol{P F}$ | $\boldsymbol{P M}$ | SWt | Sel | $\boldsymbol{C W} \boldsymbol{t}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1}$ | 4969 | 0.12 | 0 | 0 | 0 | 0.118 | 0.001 | 0.163 |
| $\mathbf{2}$ | 1736 | 0.12 | 0.26 | 0 | 0 | 0.213 | 0.131 | 0.259 |
| $\mathbf{3}$ | 3776 | 0.12 | 0.52 | 0 | 0 | 0.310 | 0.396 | 0.355 |
| $\mathbf{4}$ | 2389 | 0.12 | 0.86 | 0 | 0 | 0.408 | 0.475 | 0.453 |
| $\mathbf{5}$ | 501 | 0.12 | 1 | 0 | 0 | 0.507 | 0.444 | 0.552 |
| $\mathbf{6}$ | 232 | 0.12 | 1 | 0 | 0 | 0.608 | 0.427 | 0.652 |
| $\mathbf{7}$ | 130 | 0.12 | 1 | 0 | 0 | 0.709 | 0.434 | 0.754 |
| $\mathbf{8}$ | 64 | 0.12 | 1 | 0 | 0 | 0.812 | 0.438 | 0.856 |
| $\mathbf{9}$ | 37 | 0.12 | 1 | 0 | 0 | 0.916 | 0.434 | 0.960 |
| $\mathbf{1 0}$ | 41 | 0.12 | 1 | 0 | 0 | 1.194 | 0.434 | 1.237 |

2011

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1}$ | 4969 | 0.12 | 0 | 0 | 0 | 0.118 | 0.001 | 0.163 |
| $\mathbf{2}$ | . | 0.12 | 0.26 | 0 | 0 | 0.213 | 0.131 | 0.259 |
| $\mathbf{3}$ | . | 0.12 | 0.52 | 0 | 0 | 0.310 | 0.396 | 0.355 |
| $\mathbf{4}$ | . | 0.12 | 0.86 | 0 | 0 | 0.408 | 0.475 | 0.453 |
| $\mathbf{5}$ | . | 0.12 | 1 | 0 | 0 | 0.507 | 0.444 | 0.552 |
| $\mathbf{6}$ | . | 0.12 | 1 | 0 | 0 | 0.608 | 0.427 | 0.652 |
| $\mathbf{7}$ | . | 0.12 | 1 | 0 | 0 | 0.709 | 0.434 | 0.754 |
| $\mathbf{8}$ | . | 0.12 | 1 | 0 | 0 | 0.812 | 0.438 | 0.856 |
| $\mathbf{9}$ | . | 0.12 | 1 | 0 | 0 | 0.916 | 0.434 | 0.960 |
| $\mathbf{1 0}$ | . | 0.12 | 1 | 0 | 0 | 1.194 | 0.434 | 1.237 |

2012

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1}$ | 4969 | 0.12 | 0 | 0 | 0 | 0.118 | 0.001 | 0.163 |
| $\mathbf{2}$ | . | 0.12 | 0.26 | 0 | 0 | 0.213 | 0.131 | 0.259 |
| $\mathbf{3}$ | . | 0.12 | 0.52 | 0 | 0 | 0.310 | 0.396 | 0.355 |
| $\mathbf{4}$ | . | 0.12 | 0.86 | 0 | 0 | 0.408 | 0.475 | 0.453 |
| $\mathbf{5}$ | . | 0.12 | 1 | 0 | 0 | 0.507 | 0.444 | 0.552 |
| $\mathbf{6}$ | . | 0.12 | 1 | 0 | 0 | 0.608 | 0.427 | 0.652 |
| $\mathbf{7}$ | . | 0.12 | 1 | 0 | 0 | 0.709 | 0.434 | 0.754 |
| $\mathbf{8}$ | . | 0.12 | 1 | 0 | 0 | 0.812 | 0.438 | 0.856 |
| $\mathbf{9}$ | . | 0.12 | 1 | 0 | 0 | 0.916 | 0.434 | 0.960 |
| $\mathbf{1 0}$ | . | 0.12 | 1 | 0 | 0 | 1.194 | 0.434 | 1.237 |

Input units are thousands and kg - output in tonnes

Table 8.2.14 VIle plaice : management option table - status quo forecast

MFDP version 1a
Run: ple7e2010
WESTERN CHANNEL PLAICE,2010 WG, Forecast Inputs
Time and date: 14:07 16/05/2010
Fbar age range: 3-6

| 2010 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 3724 | 2165 | 1.0000 | 0.4356 | 1079 |


| 2011 <br> Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3947 | 2337 | 0.0000 | 0.0000 | 0 | 5371 | 3441 |
| . | 2337 | 0.1000 | 0.0436 | 124 | 5241 | 3321 |
| . | 2337 | 0.2000 | 0.0871 | 242 | 5115 | 3205 |
| . | 2337 | 0.3000 | 0.1307 | 357 | 4995 | 3095 |
| . | 2337 | 0.4000 | 0.1742 | 466 | 4879 | 2989 |
| . | 2337 | 0.5000 | 0.2178 | 572 | 4768 | 2887 |
| . | 2337 | 0.6000 | 0.2613 | 673 | 4662 | 2790 |
| . | 2337 | 0.7000 | 0.3049 | 770 | 4559 | 2697 |
| . | 2337 | 0.8000 | 0.3485 | 864 | 4461 | 2607 |
| . | 2337 | 0.9000 | 0.3920 | 954 | 4366 | 2522 |
| . | 2337 | 1.0000 | 0.4356 | 1040 | 4275 | 2439 |
| . | 2337 | 1.1000 | 0.4791 | 1124 | 4188 | 2361 |
| . | 2337 | 1.2000 | 0.5227 | 1204 | 4104 | 2285 |
| . | 2337 | 1.3000 | 0.5662 | 1281 | 4023 | 2212 |
| . | 2337 | 1.4000 | 0.6098 | 1355 | 3946 | 2143 |
| . | 2337 | 1.5000 | 0.6534 | 1426 | 3871 | 2076 |
| . | 2337 | 1.6000 | 0.6969 | 1495 | 3799 | 2012 |
| . | 2337 | 1.7000 | 0.7405 | 1561 | 3730 | 1951 |
| . | 2337 | 1.8000 | 0.7840 | 1624 | 3664 | 1892 |
| . | 2337 | 1.9000 | 0.8276 | 1686 | 3600 | 1835 |
| . | 2337 | 2.0000 | 0.8712 | 1745 | 3538 | 1781 |

Input units are thousands and kg - output in tonnes

Table 8.2.15 VIIe plaice : forecast detailed results - status quo projection

MFDP version 1a
Run: ple7e2010
Time and date: 14:07 16/05/2010
Fbar age range: 3-6

| Year: | 2010 | F multiplier: 1 | Fbar: | 0.4356 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) SSB(Jan) | SSNos(ST) | SSB(ST) |  |
| 1 | 0.0954 | 4 | 1 | 4969 | 586 | 0 | 0 | 0 | 96 |
| 2 | 0.1314 | 202 | 52 | 1736 | 370 | 451 | 96 | 451 | 96 |
| 3 | 0.3958 | 1168 | 415 | 3776 | 1171 | 1964 | 609 | 1964 | 609 |
| 4 | 0.4754 | 856 | 388 | 2389 | 975 | 2055 | 838 | 2055 | 838 |
| 5 | 0.4436 | 170 | 94 | 501 | 254 | 501 | 254 | 501 | 254 |
| 6 | 0.4274 | 76 | 50 | 232 | 141 | 232 | 141 | 232 | 141 |
| 7 | 0.4341 | 43 | 33 | 130 | 92 | 130 | 92 | 130 | 92 |
| 8 | 0.4376 | 21 | 18 | 64 | 52 | 64 | 52 | 64 | 52 |
| 9 | 0.4344 | 12 | 12 | 37 | 34 | 37 | 34 | 37 | 34 |
| 10 | 0.4344 | 14 | 17 | 41 | 49 | 41 | 49 | 41 | 49 |
| Total |  | 2567 | 1079 | 13875 | 3724 | 5474 | 2165 | 5474 | 2165 |
| Year: | 2011 | F multiplier: 1 |  | Fbar: | 0.4356 |  |  |  |  |
| Age | F |  | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) SSB(Jan) | SSNos(ST) | SSB(ST) |
| 1 | 0.0954 | 4 | 1 | 4969 | 586 | 0 | 0 | 0 | 0 |
| 2 | 0.1314 | 511 | 132 | 4403 | 938 | 1145 | 244 | 1145 | 244 |
| 3 | 0.3958 | 418 | 148 | 1350 | 419 | 702 | 218 | 702 | 218 |
| 4 | 0.4754 | 808 | 366 | 2254 | 920 | 1939 | 791 | 1939 | 791 |
| 5 | 0.4436 | 447 | 247 | 1317 | 668 | 1317 | 668 | 1317 | 668 |
| 6 | 0.4274 | 94 | 61 | 285 | 173 | 285 | 173 | 285 | 173 |
| 7 | 0.4341 | 45 | 34 | 134 | 95 | 134 | 95 | 134 | 95 |
| 8 | 0.4376 | 25 | 21 | 75 | 61 | 75 | 61 | 75 | 61 |
| 9 | 0.4344 | 12 | 12 | 37 | 34 | 37 | 34 | 37 | 34 |
| 10 | 0.4344 | 15 | 18 | 45 | 54 | 45 | 54 | 45 | 54 |
| Total |  | 2378 | 1040 | 14869 | 3947 | 5678 | 2337 | 5678 | 2337 |


| Year: | 2012 | F multiplier: 1 |  |  |  |  |  |  | Fbar: |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos(Jan) SSB(Jan) | SSNos(ST) | SSB(ST) |  |
| 1 | 0.0954 | 4 | 1 | 4969 | 586 | 0 | 0 | 0 | 0 |
| 2 | 0.1314 | 511 | 132 | 4403 | 938 | 1145 | 244 | 1145 | 244 |
| 3 | 0.3958 | 1059 | 376 | 3424 | 1062 | 1781 | 552 | 1781 | 552 |
| 4 | 0.4754 | 289 | 131 | 806 | 329 | 693 | 283 | 693 | 283 |
| 5 | 0.4436 | 421 | 233 | 1243 | 631 | 1243 | 631 | 1243 | 631 |
| 6 | 0.4274 | 247 | 161 | 750 | 456 | 750 | 456 | 750 | 456 |
| 7 | 0.4341 | 55 | 41 | 165 | 117 | 165 | 117 | 165 | 117 |
| 8 | 0.4376 | 26 | 22 | 77 | 63 | 77 | 63 | 77 | 63 |
| 9 | 0.4344 | 14 | 14 | 43 | 39 | 43 | 39 | 43 | 39 |
| 10 | 0.4344 | 16 | 19 | 47 | 56 | 47 | 56 | 47 | 56 |
| Total |  | 2643 | 1130 | 15926 | 4275 | 5943 | 2439 | 5943 | 2439 |

Input units are thousands and kg - output in tonnes

Plaice in VIle
Stock numbers of recruits and their source for recent year classes used in predictions, and the relative (\%) contributions to landings and SSB (by weight) of these year classes

| Y ear-class | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Stock No. (thousands) | 6434 | 5686 | 1962 | 4969 | 4969 |
| of 1 year-olds |  |  |  |  |  |
| Source | XSA | XSA | XSA | GM89-07 | GM89-07 |
| Status Quo F: |  |  |  |  |  |
| \% in 2010 landings | 35.9 | 38.4 | 4.8 | 0.1 | - |
| \% in 2011 | 23.8 | 35.2 | 14.2 | 12.7 | 0.1 |
| \% in 2010 SSB | 38.7 | 28.1 | 4.4 | 0.0 | - |
| \% in 2011 SSB | 28.6 | 33.8 | 9.3 | 10.4 | 0.0 |
| \% in 2012 SSB | 18.7 | 25.9 | 11.6 | 22.6 | 10.0 |

GM : geometric mean recruitment


Table 8.2.17 VIIe plaice : Yield per recruit
MFYPR version 2 a
Run: ple7e2010wg
Time and date: 16:18 16/05/2010
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 8.8433 | 6.1164 | 6.7118 | 5.7017 | 6.7118 | 5.7017 |
| 0.1000 | 0.0436 | 0.2173 | 0.1674 | 7.0353 | 4.1784 | 4.9138 | 3.7673 | 4.9138 | 3.7673 |
| 0.2000 | 0.0871 | 0.3434 | 0.2408 | 5.9873 | 3.1167 | 3.8755 | 2.7091 | 3.8755 | 2.7091 |
| 0.3000 | 0.1307 | 0.4258 | 0.2751 | 5.3035 | 2.4640 | 3.2011 | 2.0597 | 3.2011 | 2.0597 |
| 0.4000 | 0.1742 | 0.4839 | 0.2909 | 4.8221 | 2.0314 | 2.7289 | 1.6304 | 2.7289 | 1.6304 |
| 0.5000 | 0.2178 | 0.5271 | 0.2975 | 4.4649 | 1.7290 | 2.3805 | 1.3311 | 2.3805 | 1.3311 |
| 0.6000 | 0.2613 | 0.5605 | 0.2994 | 4.1893 | 1.5087 | 2.1134 | 1.1138 | 2.1134 | 1.1138 |
| 0.7000 | 0.3049 | 0.5871 | 0.2989 | 3.9700 | 1.3429 | 1.9025 | 0.9510 | 1.9025 | 0.9510 |
| 0.8000 | 0.3485 | 0.6088 | 0.2971 | 3.7914 | 1.2147 | 1.7320 | 0.8256 | 1.7320 | 0.8256 |
| 0.9000 | 0.3920 | 0.6269 | 0.2947 | 3.6431 | 1.1134 | 1.5914 | 0.7270 | 1.5914 | 0.7270 |
| 1.0000 | 0.4356 | 0.6423 | 0.2921 | 3.5178 | 1.0317 | 1.4738 | 0.6480 | 1.4738 | 0.6480 |
| 1.1000 | 0.4791 | 0.6554 | 0.2894 | 3.4106 | 0.9647 | 1.3739 | 0.5835 | 1.3739 | 0.5835 |
| 1.2000 | 0.5227 | 0.6668 | 0.2868 | 3.3178 | 0.9089 | 1.2882 | 0.5302 | 1.2882 | 0.5302 |
| 1.3000 | 0.5662 | 0.6769 | 0.2843 | 3.2365 | 0.8618 | 1.2139 | 0.4856 | 1.2139 | 0.4856 |
| 1.4000 | 0.6098 | 0.6857 | 0.2819 | 3.1647 | 0.8217 | 1.1489 | 0.4478 | 1.1489 | 0.4478 |
| 1.5000 | 0.6534 | 0.6937 | 0.2797 | 3.1009 | 0.7870 | 1.0917 | 0.4154 | 1.0917 | 0.4154 |
| 1.6000 | 0.6969 | 0.7008 | 0.2776 | 3.0436 | 0.7569 | 1.0408 | 0.3875 | 1.0408 | 0.3875 |
| 1.7000 | 0.7405 | 0.7072 | 0.2757 | 2.9920 | 0.7304 | 0.9954 | 0.3632 | 0.9954 | 0.3632 |
| 1.8000 | 0.7840 | 0.7131 | 0.2739 | 2.9451 | 0.7070 | 0.9546 | 0.3418 | 0.9546 | 0.3418 |
| 1.9000 | 0.8276 | 0.7184 | 0.2722 | 2.9024 | 0.6862 | 0.9178 | 0.3230 | 0.9178 | 0.3230 |
| 2.0000 | 0.8712 | 0.7234 | 0.2706 | 2.8633 | 0.6675 | 0.8845 | 0.3062 | 0.8845 | 0.3062 |


| Reference point | F multiplier Absolute F |  |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.4356 |
| FMax | 0.6185 | 0.2694 |
| F0.1 | 0.2865 | 0.1248 |
| F35\%SPR | 0.3127 | 0.1362 |
|  |  |  |
| Weights in kilograms |  |  |

Table 8.2.18 VIIe plaice : Summary of FMSY estimation models
Stock name
Ple V lle
Sen filename
pleviie.sen
pf, pm

Number of iterations
1000
Simulate variation in Biological parameters TRUE
SR relationship constrained
TRUE

Ricker
937/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy |
| :--- | ---: | ---: | ---: |
| Deterministic | 0.920 | 0.346 | 5528 |
| Mean | 0.858 | 0.327 | 8024 |
| 5\%ile | 0.401 | 0.195 | 3093 |
| 25\%ile | 0.555 | 0.250 | 4033 |
| 50\%ile | 0.750 | 0.312 | 5207 |
| 75\%ile | 1.013 | 0.382 | 7125 |
| 95\%ile | 1.788 | 0.526 | 16488 |
| CV | 0.534 | 0.323 | 2.063 |


| MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta |
| :--- | ---: | ---: | ---: | ---: |
| 1972 | 1.037 | 0.460 | 3.452 | 0.00019 |
| 2429 | 1.064 | 0.501 | 3.829 | 0.00021 |
| 1239 | 0.877 | 0.144 | 2.290 | 0.00006 |
| 1564 | 0.981 | 0.359 | 2.991 | 0.00015 |
| 1878 | 1.061 | 0.493 | 3.655 | 0.00020 |
| 2295 | 1.139 | 0.642 | 4.490 | 0.00027 |
| 4205 | 1.265 | 0.874 | 5.884 | 0.00036 |
| 1.565 | 0.112 | 0.431 | 0.303 | 0.43124 |

Beverton-Holt
879/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy |
| :--- | ---: | ---: | ---: |
| Deterministic | 1.059 | 0.169 | 15786 |
| Mean | 0.996 | 0.137 | 72709 |
| 5\%ile | 0.391 | 0.019 | 6456 |
| 25\%ile | 0.563 | 0.111 | 10909 |
| 50\%ile | 0.781 | 0.143 | 18120 |
| 75\%ile | 1.149 | 0.176 | 43899 |
| 95\%ile | 2.483 | 0.226 | 320659 |
| CV | 0.693 | 0.441 | 3.054 |
|  |  |  |  |
| Per recruit |  |  |  |
|  | 0.136 | 0.114 | 0.125 |
| Deterministic | 0.115 | 0.096 | 0.106 |
| Mean | 0.002 | 0.001 | 0.002 |
| 5\%ile | 0.076 | 0.063 | 0.074 |
| 25\%ile | 0.125 | 0.105 | 0.116 |
| 50\%ile | 0.161 | 0.135 | 0.147 |
| 75\%ile | 0.205 | 0.171 | 0.182 |
| 95\%ile | 0.546 | 0.548 | 0.528 |
| CV |  |  |  |


| MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta |
| ---: | ---: | ---: | ---: | ---: |
| 2732 | 0.456 | 0.985 | 11105.500 | 2795.670 |
| 5035 | 0.452 | 1.012 | 22508.660 | 9063.629 |
| 1383 | 0.112 | 0.855 | 6522.405 | 535.177 |
| 1937 | 0.296 | 0.936 | 8346.805 | 1454.855 |
| 2735 | 0.458 | 1.002 | 11050.900 | 2864.230 |
| 4261 | 0.607 | 1.079 | 17109.800 | 6048.485 |
| 10192 | 0.776 | 1.205 | 45310.970 | 20989.390 |
| 4.497 | 0.454 | 0.108 | 5.735 | 7.829 |
|  |  |  |  |  |
|  |  |  |  |  |
| Fmax | Bmsypr | MSYpr |  |  |
| 0.270 | 1.080 | 0.299 |  |  |
| 0.266 | 2.955 | 0.299 |  |  |
| 0.023 | 0.514 | 0.208 |  | 0 |
| 0.160 | 0.761 | 0.249 |  |  |
| 0.232 | 1.101 | 0.285 |  |  |
| 0.313 | 1.755 | 0.333 |  |  |
| 0.519 | 16.753 | 0.443 |  |  |
| 0.968 | 1.820 | 0.241 |  |  |

Figure 8.2.1 VIle plaice: UK(E\&W) commercial fleet LPUE and effort; and survey CPUE


Figure 8.2.2 Plaice VIIe Discards by Semester and fleet (2009)


There was no discard sampling for Misc gears in Semester 2.

Figure 8.2.2 (cont.) Plaice VIle Discards by Quarter and fleet (2009)


Figure 8.2.2 (cont.) Plaice VIle Discards by Quarter and fleet (2009)


Figure 8.2.3 : Plaice in Division VIle Length distributions of UK (England \& Wales) landings from 2000 to 2009

| 2000 | 2005 |
| :---: | :---: |
|  |  |
| 2001 | 2006 |
| 2002 | 2007 |
| 2003 | 2008 |
| 2004 | 2009 |

Figure 8.2.4 : Plaice in Division VIle Age composition of international landings 2000-2009


Figure 8.2.5 VIIe Plaice fleet log catchability residuals from the final run

E+W BEAM TRAWL SURVEY


E+W BEAM TRAWL SURVEY


UK (E+W) OTTER TRAWL


UK (E+W) OTTER TRAWL


UK(E+W) OTTER TRAWL (HISTORIC)


UK(E+W) OTTER TRAWL (HISTORIC)


Figure 8.2.5 (cont.) VIIe Plaice fleet log catchability residuals from the final run

UK (E+W) BEAM TRAWL


UK $(E+W)$ BEAM TRAWL


UK(E+W) FSP SURVEY


UK(E+W) FSP SURVEY


SEPARABLERESIDUALS


SEPARABLERESIDUALS


Figure 8.2.6 VIIe Plaice - Surba results

Tuning fleets by year-class


Tuning fleets by year


UK WECBT


Figure 8.2.7
VIle Plaice: Retrospective XSA results
(Shrinkage SE=2.5)


Note: the retrospective analysis was run without the short FSP survey

Figure 8.2.8
Plaice in Division VIle (Western Channel)


Figure 8.2.10 VIle Plaice : Yield per recruit and short term forecast results


MFYPR version 2 a
Run: ple7e2010wg
Time and date: 16:18 16/05/2010

| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(3-6) | 1.0000 | 0.4356 |
| FMax | 0.6185 | 0.2694 |
| F0.1 | 0.2865 | 0.1248 |
| F35\%.SPR | 0.3127 | 01367 |

MFDP version 1a
Run: ple7e2010
WESTERN CHANNEL PLAICE, 2010 WG, Forecast Inputs
Time and date: 14:07 16/05/2010
Fbar age range: 3-6
Input units are thousands and kg - output in tonnes

Figure 8.2.11 Plaice in VIIe. Stock-Recruitment



Figure 8.2.12. Plaice in Divisions VIIe. MSY fitted stock and recruit relationships. Left hand panels: blue line indicates the deterministic estimate; red line median and percentiles of curves with converged estimates of $\mathrm{F}_{\text {msy. }}$. Right hand panels: curves plotted from the first 100 MCMC resamples with converged $\mathrm{F}_{\text {msy }}$ estimates. The legends for each recruitment model show the number of converged values of Fmš from the $\mathbf{1 0 0 0}$ re-samples.

## Ple VIle Ricker



Figure 8.2.13. Plaice in Divisions VIIe. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Ricker stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $F_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $F_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

Ple VIle Beverton-Holt


Figure 8.2.14. Plaice in Divisions VIIe. Estimates of F reference points and equilibrium yield and SSB against fishing mortality using Beverton and Holt stock and recruitment model. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles for converged estimates of $F_{\text {msy }}$. Right hand panels: the first 100 MCMC re-samples converged $F_{\text {msy }}$ estimates. Circles show assessment estimates with the most recent year labelled.

Ple VIle - Per recruit statistics


Figure 8.2.15. Plaice in Divisions VIIe. Fitted yield-per-recruit F reference points, yield-per-recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight-at-age, natural mortality and maturity estimates and their c.v. Left hand panels: blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels: the first 100 re-samples.


Figure 8.2.16. Plaice in Divisions VIIe. MSY diagnostics.

### 8.3 Sole in Division VIIe

## Type of assessment in 2010

This stock was placed on the observational list in 2004 and has been subject to a full assessment in subsequent years. A management plan for this stock was agreed in May 2007 (Council Regulation (EC) No 509/2007).

In 2009 WKFLAT benchmarked this assessment, but failed to develop an update procedure, because it was not possible to address or even elucidate the cause of the substantial and persistent retrospective bias in F and SSB. Consequently the WG only updated data tables, performed an assessment according to previous update settings and commented on useful indicators of stock trends.

The management plan is inoperable in the absence of an analytical TAC estimate. Following a series of analyses an interim constrained model fit to the historical information was developed and is presented as a final assessment.

## ICES advice applicable to 2009

Exploitation boundaries in relation to existing management plans
The multi-annual plan implies a $20 \%$ reduction in $F$ compared to average $F(0305)$, corresponding to landings of $650 t$.

This is a $15 \%$ reduction in the TAC compared to 2008.
Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects.

Fishing mortality around $F=0.27$ can be considered as a candidate target reference point consistent with taking high long-term yields and achieving a low risk (<5\%) of depleting the productive potential. The present fishing mortality (0.42) is above the candidate reference point.

Exploitation boundaries in relation to precautionary limits
Rebuilding the stock above Bpa in just one year would require that fishing mortality is reduced by 70\%. This would correspond to landings of around 320 tonnes in 2009.

## Conclusion on exploitation boundaries

Considering that the management plan has not been evaluated by ICES, ICES advises on the basis of precautionary limits. This corresponds to a TAC of less than 320 t in 2009.

## ICES advice applicable to 2010

Precautionary reference points established in 2001 for this stock are no longer valid and there is no accepted assessment.

Survey, lpue, and the exploratory assessment suggest low stock size and high fishing mortality relative to historic estimates.

Single stock exploitation boundaries
ICES advises on the basis of exploitation boundaries in relation to precautionary considerations that fishing effort and catches should be reduced although it is not possible to determine the appropriate scale of such reductions.

## Technical consideration

Review Group comments on the 2009 assessment:

## Technical comments

- It would be useful to include some data regarding mixed species impacts in the next Report. This may be especially relevant if a precautionary TAC is used in 2009, as discarding of sole in fisheries targeting other species may increase.
- Based on the available figures (8.3.3 a-c) it appears that sampling is adequate for the combined UK fleet (486 sampled hauls), but sparse in the French trawl fleet ( 37 sampled hauls). Based on the available data, it appears that discarding may be more prevalent in the French trawl fleet, especially in the first and second quarter. The assessment may become better informed if sampling intensity is increased in the French fleet.

Conclusions
As a formal assessment was not possible, the precautionary TAC proposed by the WG seems appropriate. Lpue indices for the beam trawl fleets have declined consistently over the available time-series. In addition, the trends in F and SSB estimated in the XSA suggest that the production of the stock is declining. Therefore, a precautionary approach seems most appropriate until a suitable assessment can be conducted.

The WG has developed an interim model specification to address the requirements of the management plan.

Mixed fisheries impacts are addressed in the assessment results summary and management advice.

### 8.3.1 General

## Stock description and management units

The TAC is specified for ICES Area VIIe consistent with the assessment area.
Official national landings data as reported to ICES and the landings estimates as used by the Working Group are given in Table 8.3.1.

Official landings in 2009 were 374 t , well below the 2009 TAC ( 650 t ), but no official landings were reported by France, generally responsible for about one third of landings. WG landings included information based on French sales slips and indicated total international landings were 626 t in 2009 roughly in line with the TAC. A UK single area licence scheme introduced at the end of 2008 stopped the previous practice of misreporting; previous UK landings estimates have been corrected for area misreporting to ICES Division VIId. In previous five years landings had been stable at around 1000, with the UK taking about $65 \%$ of the TAC and France reporting the majority of the remainder.

## Management applicable to 2009 and 2010

2009

| Species. Common sole <br> Soka soka |  | Zone:VIle <br> (SOL/07E) |
| :--- | :--- | :--- |
| Belgium | 23 |  |
| France | 245 |  |
| United Kingdom | 382 | Analytical TAC <br> EC |
| TAC | 650 | Article 3 of Regulation (EC) No 847/96 <br> applie. <br> Article 4 of Regulation (EC) No 847/96 <br> applies. <br> Article 5(2) of Regulation (EC) No 847/96 <br> applies. |

In addition, Annex IIc, restricts the number of days-at-sea to 192 for beam trawlers of mesh size equal to or greater than 80 mm , and for static nets including gillnets, trammelnets and tanglenets, with mesh size less than 220 mm , with an additional 12 days for the UK beam trawl fleet due to a reduction in capacity of the fleet. In November 2008 the UK introduced a single area licence scheme to eliminate the opportunity for UK vessels to misreport catches to Area VIId.

2010

| Species: Conminu sole Sola solan |  | Zune | vile $\text { (SOL, } 07 \mathrm{E} \text { ) }$ |
| :---: | :---: | :---: | :---: |
| Belgium | 22 |  |  |
| France | 233 |  |  |
| United Kingdom | 363 |  |  |
| EU | 618 |  |  |
| IAC | 618 |  | Analytical 1 AC |

In addition, Annex IIc, restricts the number of days-at-sea to 192 for beam trawlers of mesh size equal to or greater than 80 mm , and for static nets including gillnets, trammelnets and tanglenets, with mesh size less than 220 mm , with an additional 12 days for the UK beam trawl fleet due to a reduction in capacity of the fleet. In November 2008 the UK introduced a single area licence scheme to eliminate the opportunity for UK vessels to misreport catches to Area VIId.

### 8.3.2 Data

## Landings

Total international catch numbers-at-age (Table 8.3.2, Figure 8.3.1), catch weights and stock weights-at-age (Table 8.3.3, 8.3.4, Figure 8.3.2) as used in the assessment were derived mostly by the procedure described in the Annex, except in 2009 some UK age information was used to supplement sparse French age information at larger lengths. The differences in the length distributions between the different fleets are shown in Table 8.3.5.

The UK reported landings revisions this year for 2008 (+2.0 UK; +2.0 t Guernsey).

Weights-at-age were calculated as in previous years for ages $1-14$, but the $15+$ group was not included in the within year smoothing, instead the observed weights were used, in accordance with changes with requested by WKFLAT 2009.

## Discards

Discard data suggests that discarding in 2009 is again minor in this stock (Figure 8.3.3 a-c, of both the UK and French fleets. Discarding is largely restricted to fish under the minimum landing size, although some individual discard trips may have sporadically higher rates. Discarding due to quota restrictions is rare. For a summary of historical UK discard information see the Stock Annex. In 2009 the UK provided some discard information originating from some gear experiments which indicated that discarding from the traditional gear was high at all ages. These were not included in the assessment as this discard practice was uncharacteristic of all other discard trips and the experimental conditions were deemed to have affected fishermen's behaviour.

## Biological

Natural mortality and maturity were used as in previous assessments and described in the Stock Annex.

## Survey indices

Aggregated cpue varies considerably, but indicates substantially increased catch rates from the low point of the time-series observed in 2005 to the second highest values in the time-series. (Figure 8.3.4, Table 8.3.6).

The abundance for the UK-WEC- BTS survey carried out on the chartered beam trawler FV Carhelmar is given in Table 8.3.7 and shown in Figures 8.3.4 and 8.3.5, plotted by cohort and by years. The figures show few clear year effects and good yearclass tracking for the survey at all ages until about the mid 1990s. Since then, the estimate of year-class strength at age 1 and at ages greater than 7 has deteriorated slightly. This may partly be associated with the change of vessel that occurred in 2002 and 2004 ( $R V$ Corystes used), but it seems likely this is not the only cause and weather may play a part in the catchability. One notable difference between the commercial and survey tuning-series that has been noted is the estimate of the strength of the 1998 year class. This is well represented in the commercial data, but much less clearly so in the survey data. This YC was also seen to be very strong in the VIIf\&g stock and may represent some overspill of recruitment from that stock in the adjacent western part of VIIe, not covered by the survey.

## Commercial fleets effort and lpue

Effort in hours fished for both over and under 24 m beam trawlers increased until 2001 thereafter remaining stable until 2006 (Figure 8.3.4, Table 8.3.6). Since then, $>24 \mathrm{~m}$ boats have declined in favour of smaller boats due to a combination of the UK decommissioning scheme and the substantial increases in fuel costs, making the larger boats commercially unviable. The decline of the larger boats has resulted in a resurgence of the use of vessels under 24 m . Given the licence transfer rules currently in force in the UK restructuring of the fleets will lead to a $10 \%$ decrease in the kW day capacity of replaced vessels not withstanding any latent capacity. Otter trawl effort (UK-COT) has been in continual decline since the early 1970s and is currently near the time-series low, at values roughly a third of those seen in the 1970s (Table 8.3.7). Gross registered tonnage corrected effort used in the assessment also shown in Figure 8.3.4 shows a strong decline in effort in the main fleet exploiting the stock in 2009 as
vessels moved out of the area as a result of the UK single area licensing scheme (Figure 8.3.4, Table 8.3.7).

Otter trawl effort, as used in the tuning information has been declining steadily since the late 1990s and is now at historically low levels, but takes only a small proportion of the landings.

Lpue for both over and under 24 m beam trawlers has declined steadily since 1988. Cpue from the survey is variable, but stable across this period, it is representative only of the younger ages in the fishery ( 1 to $\sim 6$ ) and only a proportion of the area exploited by the fishery.

Age disaggregated commercial abundance indices used in the assessment are the commercial beam trawl fleet (UK-CBT) and the otter trawl fleet (UK-COT) are given in Table 8.3.7, and plotted log converted by cohort and year in Figure 8.3.5 and 8.3.6 (historical fleets are retained for assessment stability). The UK-CBT shows very good year-class tracking indicated by the consistent estimation of strong and weak year classes at different ages, and demonstrates a decline in the abundance-at-age from 1975 to 1990, after which levels stabilise. There is little indication of year effects in this time-series. The UK-COT fleet also shows good year-class tracking over the middle of the time period and also gives some indication of a decline in lpue in the early 1980s although this is much less clear than in the beam trawl fleet. This is likely in part caused by the strong year effect seen for this fleet in 1991-2 and to a lesser degree in 2004. The causes of this are not clear from anecdotal evidence, but sampling for the fleet is now at relatively low levels, due to the small size of the fleet and landings.

See also the Stock Annex for historical fleet data used in the assessment.

## Information from the fishing industry

The UK fishing industry reported high abundance of sole for the area in 2009, and that improved compliance with the TAC through increased enforcement had resulted in a redistribution of effort to other divisions as well as concentrating on fishing opportunities on other species within the area.

The UK fisheries science partnership (FSP) in 2009 again conducted a survey, now in its 7th year, of sole and plaice abundance in the western channel. The results indicate that sole continue to be widespread in the area and that a large number of cohorts contribute to the stock.

### 8.3.3 Stock assessment

Model used: Reformulated XSA assessment
Software used: FLR-FLXSA (FLCore 1.4-3-"Golden Jackal"; R 2.4.1)
Model Options chosen: data as in previous years (See Stock Annex) but with additional shrinkage to stabilise F trends.

Input data types and characteristics: catch numbers-at-age without discards, five tuning fleets, one survey, two current commercial cpue series, two historical cpue series.

## Data screening

Data screening of the catch-at-age, weights, tuning information and ancillary qualitative information was carried out by the procedures set out in the Annex.

Single fleet XSA's for the current tuning fleets (see Annex for procedures) were run. Residuals for all single fleet runs were generally small (Figure 8.3.7). Residuals of the single fleet runs indicated a small but persistent decreasing trend for the CBT fleet, two large negative residuals in the COT fleet in 1992 and 2003-4 and more variable, but largely unbiased residuals for the UK-WEC-BTS. The characteristics of the individual tuning fleets are consistent with those shown previously in the screening of the tuning fleet data and hence suggest that all tuning fleets are largely consistent with the available landings data.

Summary plots of the single fleet runs are shown in Figure 8.3.8 indicate F, SSB and recruitment estimates are consistent between the fleets, although the final estimates vary slightly, with UK-CBT giving the highest F values, followed by the UK-WECBTS and the UK-COT fleets with the reverse being true of the SSB estimates.

## Final assessment

WKFLAT 2009 described the assessment methodology used prior to 2009 as unsuitable for management advice, but failed to develop a more suitable methodology. The management plan is inoperable in the absence of an analytical TAC estimate.

The WG fitted the XSA model using the previous setting, which indicated a much reduced retrospective pattern in the last two years, and considered re-introducing the old assessment methodology. However, the retrospective bias observed in previous years remained apparent and no explanations for the historical pattern can be given. Previous studies by the ICES Working Group on Assessment Methods (ICES 1991, ICES CM1991/ Assess:25) established that where retrospective bias patterns are severe, such that estimates are considered unreliable, shrinkage to the mean fishing mortality of the previous years at each age could be used to provide coherent population and fishing mortality estimates that can be taken forward into stock forecasts. An XSA with heavy shrinkage was therefore considered to be the most likely methodology to provide quantitative information suitable for management advice. The results from exploratory runs established that an increased level of shrinkage ( 0.5 from 1.0) and an increased time period over which this is applied (10 years from 5 years) was optimal for consistent series of estimates. All other settings were maintained as previously and the complete set of settings is shown in the text table below.

Figure 8.3.9 shows the residual plots from the final fitted model. Figure 8.3.10 shows a comparison with the 2009 assessment. XSA diagnostic tables, fishing mortality-atage, and stock numbers-at-age are shown in Tables 8.3.8-8.3.10. Comparisons of the XSA estimate weighting between the previous, SPALY and current assessment highlights the increased weighting given to F-shrinkage in this year's assessment in order to decrease the retrospective bias (Figure 8.3.11).

A seven year retrospective analysis was run for the interim assessment (Figure 8.3.12), which still shows some retrospective bias prior to 2006 , but confirms that the more recent period is more stable with respect to F and SSB trends.


## State of the stock

Stock trends are shown in Table 8.3.11 and plotted in Figure 8.3.10.
SSB is estimated to have increased from 1970 to 1980 following successive strong recruitments. Subsequently it has declined until 1993 after which it remained stable for 12 years before declining slightly to historically low levels in 2006/2007. There has been a slight increase in the last two years.

The base level of recruitment has remained stable during the whole time-series in the range $4-5$ million recruits. The main development has been a reduction in recruitment variability since 1991 with very few if any of the very abundant year classes that maintained a higher level of biomass during the early period.

Fishing mortality was stable at a low level until 1977 after which it increased sharply until 1982, remained relatively constant until 2004 and then increased until 2007. F decreased slightly in 2008 and then sharply in 2009, commensurate with the improved compliance associated with the single area licensing scheme introduced in the UK.

Information that is consistent with the decrease in fishing mortality in the most recent year is provided by the decline in effort (Figure 8.3.4) and landings (Table 8.3.1) which have decreased to levels close to the TAC.

The age structure of the VIIe sole stock continues to be more extended than other sole stocks in European waters, implying low mortality rates, with the plus group (at age 12) containing a high proportion of the catches and including some individual of ages 33-38 in recent years.

### 8.3.4 Short-term projections

A short-term forecast based on Fsq has been used previously for this stock, because area misreporting had meant that in the past the TAC had not been constraining. However, recent evidence suggests that the TAC is likely to be observed so a TAC constraint forecast was implemented this year using the average stock and catch weights 07-09 and the selectivity pattern 07-09 scaled to achieve a 2010 TAC of 618 t for the interim year.

## Estimating year class abundance

As implemented previously, the geometric mean recruitment over the entire timeseries (69-07) was used as there is no evidence of a significant relationship between SSB and subsequent recruitment over the range of SSB values observed in the assessment.

| Year class | Thousands | Basis | Surveys | Commercial | Shrinkage |
| ---: | :---: | :--- | :---: | :---: | :---: |
| 2007 | 3037 | XSA | $62 \%$ | - | $38 \%$ |
| 2008 | 4332 | GM $(69-07)$ |  |  |  |
| 2009 | 4332 | GM $(69-07)$ |  |  |  |
| 2010 | 4332 | GM $(69-07)$ |  |  |  |

Complete input data for the short-term forecast is shown in Table 8.3.12, and resulting $\mathrm{F}_{\mathrm{sq}}$ forecast constrained in 2010 to the 618 t TAC is shown in Table 8.3.13.

SSB estimated at 2470 t in 2009 will remain stable at 2400 t in 2010, despite a decrease in F in $2009(0.25)$ as a result of weaker recruitment $(07,08)$ but will increase at $\mathrm{F}_{\text {sq }}$ (0.355) in 2011 to 2540 t as cohorts for which $\mathrm{GM}_{(69-07)}$ has been assumed enter the fishery.

The proportions that the 2007-2011 year classes will contribute to the landings in 2011, and to the SSB in 2012, are given in Table 8.3.14. 29\% of the landings for 2011 and $40 \%$ of the SSB for 2012 rely on year classes for which GM recruitment has been assumed. The 2008 year class that has been replaced with GM (69-07) contributes $21 \%$ to the landings in 2010 and $23 \%$ of the SSB in 2012.

A full management options table is provided in Table 8.3.15. The management plan for this stock requires a reduction in F of $85 \%$ of the average $\mathrm{F}_{07-09}=0.36$ ( $\mathrm{F}_{2011}=0.3$ ) resulting in a yield of 809 t in 2011 and representing a $31 \%$ increase in the yield over 2010. The TAC-stability clause in the management plan allowing for a maximum $15 \%$ change in TAC from year to year will limit the TAC to 710 t equivalent to an $\mathrm{F}=0.25$ just below the long-term management target of $\mathrm{F}_{\text {manag }}=0.27$.

### 8.3.5 Biological reference points

Biological reference points were rejected by WKFLAT 2009 due to a lack of an appropriate assessment to evaluate their suitability. Reference points should be revised once an appropriate assessment methodology has been fully developed.

ICES is moving towards management advice based on MSY reference points. Management of this stock in the short term is likely to be on the basis of the management plan so that a lack of PA reference points is unlikely to significantly impact management decisions in either the short or medium term.

### 8.3.6 MSY evaluation

Although there is no benchmark procedure for providing a yield-per-recruit or MSY analysis for this stock, ICES ACOM requires the provision of such information to inform the commission on suitable long-term management targets. The interim assessment provided by the WG is reasonably stable with regards to estimates of selectivity, especially when using the average selectivity pattern over the last seven years. Input data used by the SRMSYMC program is shown in the .sen file (Table 8.3.16). A thousand MCMC iterations were provided, roughly $90 \%$ of which provided converged $\mathrm{F}_{\text {msy }}$ estimates for the various stock-recruit relationships investigated.

Stock-recruitment plots for the three stock-recruitment models are shown in Figure 8.3.12 and indicate a cloud of stock-recruit points through which it is possible to draw plausible fits of any of the stock-recruitment relationships, with little information to distinguish the suitability of any of them.

As noted previously, the base level of recruitment has remained stable during the whole time-series in the range $4-5$ million recruits. The main development has been a reduction in recruitment variability since 1991 with none of the very large year classes that maintained a higher level of biomass during the early period. The model fits pass through the cloud of estimates and there is no information as to the structure of the relationship (Table 8.3.16).

All models imply that the population in recent years has been exploited at levels close to or above Fcrash. The stock has been exploited at these levels since the 1980s, and SSB and recruitment have been stable for the last 20 years. Based on the unrealistic estimates of $\mathrm{F}_{\text {crash }}$ all stock-recruit models were rejected.

The yield-per-recruit analysis indicates that all reference points are poorly defined when low levels of uncertainty are associated with natural mortality, maturity and selection and weights-at-age.

Higher yields and lower risk to the stock can be obtained at levels of F lower than those observed in the last 10 years. However, little or no information is available on the stock dynamics at such levels so that it is not possible to provide estimates of $\mathrm{F}_{\text {msy }}$ for use as management targets based on the analysis performed during this WG.

Stochastic analyses performed by WGSSDS in 2006, assuming no variability in M, suggested that yields of 865 t could be safely extracted from the stock at levels of $\mathrm{F}=0.27$ while the probability of SSB dropping below lowest observed SSB values would remain at less than $5 \%$. This value remains consistent with the results of the current assessment and is accepted as the best estimate of $\mathrm{F}_{\mathrm{msy}}$ available for this stock.

### 8.3.7 Management plan

The commission implemented a management plan for the recovery of the stock early in 2007 (Council Regulation (EC) No 509/2007). ICES evaluated the management plan and concluded that:

The long-term management target $(\mathrm{F}=0.27)$ is precautionary in the sense that it ensures that there is a less than $5 \%$ chance of SSB declining below previously observed levels, as well as maintaining yield within $10 \%$ of MSY (WGCSE note: long term yield at $F_{\max }$ ) (WG 2005, WG 2006).

The methodology of reaching the long-term target in 3-year stepped reductions in F is also acceptable. However, the size of further steps is based on observed fishing mortalities within the period of the management plan. This
can only have the desired effect if management measures (TAC) are effective and if estimates of recent levels of F from the assessment are accurate. In 2009 newly introduced enforcement measures appear to have resulted in increased compliance with the TAC; continued development of the SSB will be dependent on effective controls of fishing effort.

The WG has provided an interim assessment as a means of providing management advice for 2011, which implies a TAC of 700 t in 2011. Catches at this level are likely to maintain F around the long-term management target.

### 8.3.8 Uncertainties in assessment and forecast

The WG provided a constrained interim assessment due to the need for management advice in relation to the management plan, as the current plan makes no provisions as to how to manage the stock in the absence of a full analytical assessment. The methodology provided is as robust as possible under the currently available understanding of the stock dynamics and at present does not appear to suffer from the retrospective pattern, which led to the rejection of the assessment as suitable for management advice by WKFLAT 2009. However, the retrospective analysis suggest that even the new methodology still retains some retrospective bias in the earlier period so that the uncertainty in the current estimates of F and SSB is likely to be greater than indicated by the assessment output diagnostics. The absolute values are used in the short-term forecast and the uncertainty is used in the MSY analysis.

## Sampling

Age and length sampling for this stock is mostly adequate. Age data from the largest two sectors prosecuting this fishery (UK and France, together about $95 \%$ of landings) are included in the assessment. French age data in 2009 was insufficient at older ages to raise the length compositions, so that UK data was used to cover the larger fish. The use of commercial tuning data is unavoidable, as there is little information available for older ages from the survey.

## Discarding

There is currently little discarding of undersized fish in this stock. There is some anecdotal evidence of a change in high-grading practice in the UK fleet due to the introduction of a single area licensing scheme. Despite this, high-grading was not found to be a significant contributor to mortality in 2009 in the UK discard data. However, gear trials conducted by the UK suggested that incentives to high-grade remain under the current fleet capacity. If this is found to be significant in future it should be possible to include this in the assessment in subsequent years, as it would not require a time-series of discards unlike most other stocks currently assessed without discards.

## Surveys

Currently only one survey index is used in the assessment (UK-WEC-BTS) which provides stability to the assessment in general. Year-class tracking is internally consistent and agrees reasonably with information from commercial tuning fleets. However, in the recent past there is some question regarding the consistency of the tuning-series due to a vessel effect in 2002 and 2004. In addition in recent years it has become apparent that there are some differences in the year-class consistency between the commercial and survey tuning information. Specifically, the 1998 year class known to have been very strong in VIIf\&g is not represented in the survey that oper-
ates solely in the eastern part of the area. This suggests that there may be both an open population as well as an incomplete mixing problem in the data contributing to the inconsistency of the assessment.

## Consistency

The interim assessment provided by the WG is mostly consistent with the previous methodology as it uses the same information, but weights the value of the different sources differently. The estimates of stock status in 2009 are consistent with those produced by the previous assessment methodology.

## Misreporting

Area misreporting, mainly to Area VIId had declined to low levels in recent years, through a combination of enforcement and a substantial increase in the TAC in 2005. There have also been some attempts to prosecute UK fishermen for misreporting to Area VIIh, although to date none of those prosecutions have been successful for lack of legally acceptable evidence.

Levels of underreporting are thought to have been serious in the early 1980s prior to the shift to area misreporting. Although it is clear that levels of underreporting are also much lower now, no quantitative information is available on the size of the problem.

Landings in 2009 were in line with the TAC for the first time, suggesting improved compliance. The decrease in landings is also consistent with a reduction in effort by the main fleet and a reduction in F observed in the plaice VIIe stock, a major bycatch of the sole fishery.

### 8.3.9 Recommendation for the next Benchmark

| Year | Candidate Stock | Supporting Justification | Suggested time | Indicate expertise necessary at benchmark meeting. |
| :---: | :---: | :---: | :---: | :---: |
| 2009 | VIIe Sole | WKFLAT 2009 could not recommend an appropriate assessment procedure for this stock for the following reasons: <br> Closed population and complete mixing assumptions of the assessment are violated. Tuning data indicate differences in trends in $F$ and recruitment resulting in a serious retrospective pattern in the assessment. <br> Survey information only partially covers the stock area. <br> Effort correction parameters/methodologies require updating as the main beamtrawl fleet has restructured substantially recently. <br> This effort would be greatly enhanced by an internationally coordinated survey that more appropriately covers the management area and is able to assess recruitment dynamics irrespective of the sources of recruitment and environmental drivers. | 2012 | Experts with expertise in spatial modelling of stock dynamics, expertise in the analysis of tagging information. |

### 8.3.10 Management considerations

This stock is subject to a management plan based on reductions in fishing mortality in relation to historical levels of F. Previously both the most recent and the target fishing mortality and population estimates were continually revised by subsequent assessments, which is why the assessment was rejected by WKFLAT 2009.

A constrained interim assessment model has been fitted in order to provide management advice in relation to the management plan, as the current plan makes no provisions as to how to manage the stock in the absence of a full analytical assessment. The model is considered to provide population and mortality estimates that are coherent and suitable for the provision of stock forecasts.

The management of the stock by TAC in the past has been ineffective, as the restrictiveness of the TAC was inversely related to the degree of area misreporting. In November 2008 the UK introduced a single area licensing scheme for beam trawlers in VIIe, which appears to have restricted the possibility of area misreporting in 2009 but had little or no effect in 2008.

Effort restrictions have not been sufficient to ensure an observable decrease in F in recent years. Decommissioning in the UK fleet in 2007-2008 did not reduce fleet capacity sufficiently, but TAC restrictions appear to have been effective in 2009 and resulted in the fleet utilising fishing opportunities in other ICES divisions so that effective effort in Division VIIe dropped markedly.

Plaice are taken as a bycatch in this fishery, so that management advice for sole must also take into account the advice for plaice. The effort reductions in 2009 have also positively impacted the plaice stock with a sizeable reduction in F indicated for that stock also. Angler fish, cuttle fish, and lemon sole are also important bycatches in this fishery. The UK beam trawl fleet has recently started to land sizeable quantities of gurnards for human consumption.

Estimates of $\mathrm{F}_{\mathrm{msy}}$ and its proxies were all considered highly uncertain for this stock and therefore not considered appropriate. The current management plan is considered appropriate to achieving high long-term yields consistent with MSY.

### 8.3.11 Ecosystem considerations

Beam trawling, especially using chain-mat gear, is known to have a significant impact on the benthic communities, although less so on soft substrates and in areas which have been historically exploited by this fishing method. Discard rates of noncommercial species and commercial species of unmarketable size are substantial, but total discards are lower compared to some other gears due to the relatively small area swept by the gear.

### 8.3.12 Regulations and their effects

In November of 2008 the UK introduced a single area licensing scheme for beam trawlers, which is thought to be highly effective in eliminating the current practice of area misreporting by this fleet, but will have had little effect on the fishery in 2008. Landings and effort data for 2009 indicate that the measure has been effective.

Management of this stock is mainly by TAC. In 2005 effort restrictions were implemented for beam trawlers and entangling gears targeting sole this fishery to enforce the TAC and improve data quality. To date these restrictions have not been limiting in this fishery, in part due to the large numbers of days available, but also because in the UK fleet there appears to remain some latent effort/overcapacity in the beam trawl fleet despite decommissioning.

Mesh restrictions for towed gears are set to 80 mm codends, which correspond well with the minimum landing size of sole at 24 cm . Consequently there is little discarding of sole in this fishery.

### 8.3.13 Changes in fishing technology and fishing patterns

The UK industry has applied for MSC certification in 2009 commensurate with which it has started to adopt larger codend meshes and square mesh panels to limit the impact on benthic ecosystems. However these changes have not been adopted fully in 2008 so the effects are likely to be minimal for this year's assessment, but significant changes may be observed in the 2009 data and the WG will endeavour to monitor the situation.

### 8.3.14 Changes in the environment

WGRED 2008 overall indicated that there were no consistent environmental drivers altering the ecosystem in Celtic Sea Area, although it did provide some more detailed description of the environmental changes occurring in the system, including climate change, NAO and changes in plankton productivity and species composition.

The winter NAO experienced a strong negative phase in the 1960s, becoming more positive in the 1980s and early 1990s. It remained mainly negative from 1996 to 2004, but became positive in 2005_(6.7 mbar).

Although the assessment only goes back to 1969, relative year class for sole VIIe from catches indicates some very strong recruitment for example in 1963, following which recruitment appears to have declined coinciding with the strong negative phase of the NAO. Positive NAOs in the 1980s and 1990s coincide with some of the highest recruitments seen in the assessment, which have declined since then along with NAO values. Since 2005 the NAO again shows more favourable conditions although this has not immediately resulted in returns of the occasional exceptional year classes.

This should be investigated further by the next Benchmark.

Table 8.3.1 Sole VIIE Nominal landings ( t ) as used by the WG

| Year | Belgium | Denmark | France | Netherlands | Ireland | Jersey | Guernsey | $\begin{array}{r} \hline \text { UK E W } \\ \mathrm{Ni} \end{array}$ | $\begin{array}{r} \text { UK } \\ \text { other } \end{array}$ | Unallocated | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 |  |  | 323 |  |  |  |  |  |  | 104 | 427 |
| 1975 | 3 |  | 271 |  |  |  | 2 | 215 | 2 |  | 491 |
| 1976 | 4 |  | 352 |  |  |  | 1 | 259 | 1 |  | 616 |
| 1977 | 3 |  | 331 |  |  |  |  | 272 |  |  | 606 |
| 1978 | 4 |  | 384 |  |  |  |  | 453 |  | 20 | 861 |
| 1979 | 1 |  | 515 |  |  |  | 2 | 663 | 2 |  | 1181 |
| 1980 | 45 |  | 447 |  | 13 |  | 1 | 763 | 1 |  | 1269 |
| 1981 | 16 |  | 415 | 1 |  |  | 4 | 784 | 4 | -5 | 1215 |
| 1982 | 98 |  | 321 |  |  |  | 15 | 1013 | 15 | -1 | 1446 |
| 1983 | 47 |  | 405 | 3 |  | 2 | 16 | 1025 | 18 |  | 1498 |
| 1984 | 48 |  | 421 |  |  | 9 | 14 | 878 | 23 |  | 1370 |
| 1985 | 58 |  | 130 |  |  | 9 | 8 | 894 | 17 | 310 | 1409 |
| 1986 | 62 |  | 467 |  |  | 3 | 6 | 831 | 9 | 50 | 1419 |
| 1987 | 48 |  | 432 |  |  | 1 | 5 | 626 | 6 | 168 | 1280 |
| 1988 | 67 |  | 98 |  |  | 0 | 4 | 780 | 4 | 495 | 1444 |
| 1989 | 69 |  | 112 | 6 |  |  | 3 | 610 | 3 | 590 | 1390 |
| 1990 | 41 | 0 | 81 |  |  | 1 | 3 | 632 | 4 | 556 | 1315 |
| 1991 | 35 |  | 325 |  |  |  |  | 477 |  | 15 | 852 |
| 1992 | 41 |  | 267 |  |  |  | 2 | 457 | 11 | 119 | 895 |
| 1993 | 59 |  | 236 |  |  | 1 |  | 479 | 19 | 111 | 904 |
| 1994 | 33 |  | 257 |  |  |  |  | 546 | 2 | -38 | 800 |
| 1995 | 21 |  | 294 |  |  | 1 | 2 | 562 | 3 | -24 | 856 |
| 1996 | 8 |  | 297 |  |  |  |  | 428 | 9 | 91 | 833 |
| 1997 | 13 |  | 348 |  | 1 | 13 | 13 | 470 | 26 | 91 | 949 |
| 1998 | 40 |  | 343 |  |  | 17 | 3 | 369 | 20 | 108 | 880 |
| 1999 | 13 |  |  |  |  | 18 | 3 | 375 | 21 | 548 | 957 |
| 2000 | 4 |  | 241 |  |  | 22 | 5 | 386 | 27 | 256 | 914 |
| 2001 | 19 |  | 224 |  |  | 20 | 5 | 382 | 25 | 419 | 1069 |
| 2002 | 33 |  | 198 |  |  | 15 | 5 | 289 | 20 | 566 | 1106 |
| 2003 | 1 |  | 363 |  | 1 | 15 | 5 | 235 | 20 | 458 | 1078 |
| 2004 | 7 |  | 302 |  |  | 7 | 6 | 172 | 13 | 581 | 1075 |
| 2005 | 26 |  | 406 |  |  | 17 | 5 | 505 | 22 | 80 | 1039 |
| 2006 | 32 |  | 357 |  |  | 4 | 4 | 568 | 8 | 57 | 1022 |
| 2007 | 34 |  | 383 |  | 2 | 2 |  | 525 | 2 | 69 | 1015 |
| 2008 | 28 |  | 183 |  | 0 | 2 | 6 | 463 | 8 | 230 | 908 |
| 2009 | 18 |  |  |  | 1 | 1 |  | 354 | 1 | 252 | 626 |

Table 8.3.2 Sole VIIE Catch Numbers at Age in 000's

| Age | 1969 |  | 1970 | 1971 |
| ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 |  |
| 2 | 89 | 53 | 51 |  |
| 3 | 322 | 232 | 200 |  |
| 4 | 80 | 322 | 246 |  |
| 5 | 148 | 90 | 198 |  |
| 6 | 210 | 83 | 65 |  |
| 7 | 21 | 112 | 80 |  |
| 8 | 50 | 13 | 156 |  |
| 9 | 26 | 35 | 10 |  |
| 10 | 20 | 52 | 35 |  |
| 11 | 9 | 22 | 54 |  |
| + gp | 63 | 113 | 113 |  |
| Total | 1037 | 1127 | 1207 |  |
| Landings | 353 | 391 | 432 |  |

Table 8.3.2 Sole VIIE Catch Numbers at Age in 000's continued

| Age | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 146 | 71 | 45 | 82 | 167 | 426 | 250 | 227 | 175 | 245 |
| 3 | 412 | 396 | 349 | 567 | 419 | 318 | 1123 | 803 | 559 | 806 |
| 4 | 167 | 433 | 220 | 170 | 472 | 384 | 347 | 811 | 497 | 651 |
| 5 | 115 | 89 | 178 | 199 | 161 | 206 | 214 | 250 | 630 | 467 |
| 6 | 112 | 99 | 71 | 115 | 135 | 102 | 189 | 229 | 126 | 389 |
| 7 | 14 | 120 | 80 | 28 | 92 | 70 | 103 | 174 | 183 | 179 |
| 8 | 25 | 17 | 43 | 53 | 46 | 74 | 72 | 103 | 140 | 126 |
| 9 | 134 | 52 | 32 | 26 | 58 | 10 | 77 | 90 | 65 | 76 |
| 10 | 38 | 30 | 24 | 22 | 51 | 24 | 38 | 104 | 56 | 58 |
| 11 | 54 | 4 | 55 | 24 | 14 | 32 | 27 | 28 | 130 | 55 |
| + gp | 106 | 136 | 106 | 171 | 213 | 159 | 203 | 290 | 342 | 211 |
| Total | 1323 | 1446 | 1202 | 1456 | 1830 | 1804 | 2644 | 3108 | 2902 | 3262 |
| Landings | 437 | 459 | 427 | 491 | 616 | 606 | 861 | 1181 | 1269 | 1215 |

Table 8.3.2 Sole VIIE Catch Numbers at Age in 000's continued

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 128 | 91 | 333 | 287 | 246 | 487 | 443 | 390 | 341 | 450 |
| 3 | 1451 | 753 | 663 | 1700 | 1618 | 808 | 1438 | 871 | 902 | 415 |
| 4 | 916 | 1573 | 826 | 756 | 971 | 1090 | 596 | 1233 | 581 | 482 |
| 5 | 553 | 583 | 758 | 469 | 421 | 427 | 728 | 497 | 553 | 289 |
| 6 | 352 | 351 | 325 | 585 | 321 | 204 | 374 | 509 | 244 | 220 |
| 7 | 240 | 267 | 204 | 179 | 336 | 224 | 153 | 225 | 264 | 93 |
| 8 | 136 | 294 | 129 | 97 | 84 | 229 | 162 | 110 | 143 | 111 |
| 9 | 113 | 119 | 152 | 103 | 75 | 47 | 109 | 107 | 103 | 68 |
| 10 | 81 | 73 | 54 | 85 | 90 | 50 | 39 | 113 | 75 | 37 |
| 11 | 61 | 37 | 28 | 29 | 74 | 41 | 50 | 48 | 85 | 31 |
| +gp | 294 | 262 | 255 | 125 | 127 | 162 | 171 | 214 | 235 | 145 |
| Total | 4324 | 4401 | 3727 | 4414 | 4363 | 3770 | 4262 | 4316 | 3525 | 2341 |
| Landings | 1446 | 1498 | 1370 | 1409 | 1419 | 1280 | 1444 | 1390 | 1315 | 852 |

Table 8.3.2 Sole VIIE Catch Numbers at Age in 000's continued

| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 316 | 209 | 97 | 95 | 365 | 216 | 265 | 280 | 307 | 145 |
| 3 | 1434 | 704 | 657 | 308 | 445 | 831 | 606 | 915 | 599 | 1401 |
| 4 | 417 | 1107 | 558 | 629 | 364 | 724 | 536 | 500 | 751 | 531 |
| 5 | 297 | 350 | 558 | 427 | 298 | 325 | 336 | 398 | 367 | 497 |
| 6 | 115 | 219 | 112 | 411 | 235 | 180 | 209 | 255 | 229 | 268 |
| 7 | 112 | 151 | 106 | 131 | 257 | 194 | 151 | 114 | 107 | 178 |
| 8 | 61 | 78 | 49 | 101 | 68 | 173 | 80 | 103 | 53 | 100 |
| 9 | 74 | 60 | 57 | 61 | 61 | 44 | 127 | 54 | 68 | 55 |
| 10 | 26 | 56 | 44 | 33 | 49 | 20 | 35 | 107 | 51 | 43 |
| 11 | 23 | 31 | 50 | 18 | 37 | 40 | 34 | 25 | 88 | 42 |
| +gp | 90 | 79 | 99 | 142 | 143 | 88 | 162 | 123 | 91 | 159 |
| Total | 2964 | 3045 | 2388 | 2356 | 2321 | 2835 | 2543 | 2874 | 2710 | 3419 |
| Landings | 895 | 904 | 800 | 856 | 833 | 949 | 880 | 957 | 914 | 1069 |

Table 8.3.2 Sole VIIE Catch Numbers at Age in 000's continued

| Age | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | geom <br> mean | arith <br> mean <br> $07-09$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  | $07-09$ |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |
| 2 | 332 | 598 | 398 | 258 | 500 | 201 | 281 | 122 | 190.38 | 201.36 |
| 3 | 1251 | 835 | 1080 | 468 | 786 | 852 | 752 | 393 | 631.50 | 665.69 |
| 4 | 843 | 953 | 448 | 834 | 472 | 755 | 678 | 339 | 557.79 | 590.66 |
| 5 | 387 | 645 | 445 | 449 | 606 | 293 | 376 | 302 | 321.56 | 323.58 |
| 6 | 322 | 130 | 526 | 366 | 250 | 362 | 163 | 188 | 222.94 | 237.61 |
| 7 | 129 | 74 | 164 | 293 | 224 | 179 | 184 | 62 | 126.75 | 141.56 |
| 8 | 105 | 50 | 116 | 113 | 185 | 130 | 105 | 67 | 96.92 | 100.52 |
| 9 | 94 | 58 | 61 | 80 | 85 | 110 | 71 | 35 | 64.74 | 71.94 |
| 10 | 33 | 63 | 54 | 45 | 56 | 55 | 67 | 46 | 55.61 | 56.29 |
| 11 | 18 | 14 | 35 | 24 | 31 | 27 | 39 | 32 | 32.16 | 32.57 |
| + gp | 85 | 61 | 85 | 96 | 87 | 99 | 89 | 59 | 80.29 | 82.25 |
| Total | 3599 | 3482 | 3412 | 3027 | 3282 | 3062 | 2805 | 1645 | 2417.49 | 2504.04 |
| Landings | 1106 | 1078 | 1075 | 1039 | 1023 | 1015 | 908 | 626 | 832.48 | 849.67 |

Table 8.3.3 Sole VIIE Catch Weights at Age in kgs

| Age | 1969 | 1970 |
| ---: | ---: | ---: |
| 1 | 0.000 | 0.000 |
| 2 | 0.188 | 0.187 |
| 3 | 0.245 | 0.223 |
| 4 | 0.332 | 0.294 |
| 5 | 0.329 | 0.314 |
| 6 | 0.367 | 0.354 |
| 7 | 0.522 | 0.434 |
| 8 | 0.455 | 0.498 |
| 9 | 0.463 | 0.442 |
| 10 | 0.606 | 0.512 |
| 11 | 0.647 | 0.528 |
| +gp | 0.660 | 0.593 |

Table 8.3.3 Sole VIIE Catch Weights at Age in kgs continued

| Age | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.113 | 0.000 | 0.000 | 0.144 | 0.142 | 0.139 | 0.118 | 0.000 | 0.000 | 0.000 |
| 2 | 0.151 | 0.194 | 0.203 | 0.183 | 0.181 | 0.170 | 0.197 | 0.180 | 0.187 | 0.189 |
| 3 | 0.222 | 0.227 | 0.224 | 0.224 | 0.214 | 0.217 | 0.248 | 0.241 | 0.237 | 0.254 |
| 4 | 0.296 | 0.272 | 0.262 | 0.281 | 0.299 | 0.286 | 0.302 | 0.303 | 0.327 | 0.343 |
| 5 | 0.367 | 0.369 | 0.310 | 0.379 | 0.358 | 0.323 | 0.356 | 0.390 | 0.423 | 0.389 |
| 6 | 0.350 | 0.408 | 0.381 | 0.434 | 0.403 | 0.390 | 0.399 | 0.439 | 0.460 | 0.525 |
| 7 | 0.359 | 0.458 | 0.414 | 0.372 | 0.435 | 0.454 | 0.502 | 0.377 | 0.468 | 0.560 |
| 8 | 0.431 | 0.495 | 0.459 | 0.464 | 0.497 | 0.413 | 0.463 | 0.486 | 0.477 | 0.609 |
| 9 | 0.455 | 0.402 | 0.466 | 0.475 | 0.591 | 0.475 | 0.517 | 0.489 | 0.565 | 0.646 |
| 10 | 0.476 | 0.454 | 0.537 | 0.487 | 0.651 | 0.478 | 0.484 | 0.488 | 0.522 | 0.655 |
| 11 | 0.388 | 0.508 | 0.654 | 0.474 | 0.535 | 0.583 | 0.552 | 0.540 | 0.569 | 0.600 |
| +gp | 0.653 | 0.600 | 0.561 | 0.731 | 0.676 | 0.628 | 0.681 | 0.670 | 0.725 | 0.783 |

Table 8.3.3 Sole VIIE Catch Weights at Age in kgs continued

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.000 | 0.120 | 0.000 | 0.088 | 0.000 | 0.106 | 0.098 | 0.091 | 0.110 | 0.158 |
| 2 | 0.174 | 0.213 | 0.188 | 0.209 | 0.162 | 0.174 | 0.174 | 0.170 | 0.167 | 0.216 |
| 3 | 0.226 | 0.208 | 0.251 | 0.242 | 0.225 | 0.237 | 0.245 | 0.244 | 0.222 | 0.270 |
| 4 | 0.322 | 0.276 | 0.272 | 0.304 | 0.296 | 0.297 | 0.310 | 0.312 | 0.275 | 0.322 |
| 5 | 0.382 | 0.345 | 0.307 | 0.379 | 0.358 | 0.354 | 0.370 | 0.375 | 0.326 | 0.370 |
| 6 | 0.478 | 0.424 | 0.390 | 0.389 | 0.389 | 0.407 | 0.425 | 0.432 | 0.375 | 0.416 |
| 7 | 0.515 | 0.495 | 0.419 | 0.478 | 0.469 | 0.456 | 0.474 | 0.484 | 0.422 | 0.458 |
| 8 | 0.534 | 0.507 | 0.475 | 0.539 | 0.520 | 0.502 | 0.518 | 0.531 | 0.467 | 0.498 |
| 9 | 0.599 | 0.520 | 0.532 | 0.559 | 0.531 | 0.544 | 0.557 | 0.572 | 0.510 | 0.534 |
| 10 | 0.620 | 0.523 | 0.610 | 0.601 | 0.519 | 0.583 | 0.590 | 0.608 | 0.551 | 0.567 |
| 11 | 0.710 | 0.561 | 0.553 | 0.722 | 0.584 | 0.618 | 0.618 | 0.639 | 0.590 | 0.597 |
| +gp | 0.661 | 0.659 | 0.667 | 0.639 | 0.817 | 0.703 | 0.665 | 0.694 | 0.692 | 0.664 |

Table 8.3.3 Sole VIIE Catch Weights at Age in kgs continued

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.105 | 0.088 | 0.000 | 0.122 | 0.133 | 0.164 | 0.000 | 0.000 | 0.158 | 0.141 |
| 2 | 0.182 | 0.166 | 0.146 | 0.183 | 0.192 | 0.214 | 0.186 | 0.191 | 0.208 | 0.201 |
| 3 | 0.255 | 0.238 | 0.209 | 0.241 | 0.248 | 0.262 | 0.244 | 0.247 | 0.257 | 0.257 |
| 4 | 0.323 | 0.305 | 0.268 | 0.295 | 0.301 | 0.308 | 0.300 | 0.300 | 0.303 | 0.309 |
| 5 | 0.386 | 0.366 | 0.324 | 0.347 | 0.351 | 0.354 | 0.354 | 0.350 | 0.347 | 0.357 |
| 6 | 0.445 | 0.423 | 0.376 | 0.396 | 0.397 | 0.399 | 0.406 | 0.397 | 0.389 | 0.400 |
| 7 | 0.499 | 0.474 | 0.425 | 0.442 | 0.441 | 0.442 | 0.455 | 0.441 | 0.429 | 0.440 |
| 8 | 0.549 | 0.520 | 0.470 | 0.484 | 0.481 | 0.484 | 0.503 | 0.482 | 0.467 | 0.475 |
| 9 | 0.594 | 0.561 | 0.513 | 0.524 | 0.518 | 0.524 | 0.548 | 0.520 | 0.502 | 0.507 |
| 10 | 0.634 | 0.597 | 0.551 | 0.561 | 0.552 | 0.564 | 0.592 | 0.555 | 0.535 | 0.534 |
| 11 | 0.669 | 0.627 | 0.587 | 0.595 | 0.583 | 0.602 | 0.633 | 0.586 | 0.566 | 0.557 |
| +gp | 0.742 | 0.684 | 0.672 | 0.671 | 0.652 | 0.695 | 0.734 | 0.661 | 0.636 | 0.645 |

Table 8.3.3 Sole VIIE Catch Weights at Age in kgs continued

| Age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | mean <br> $07-09$ |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.000 | 0.123 | 0.101 | 0.122 | 0.123 | 0.106 | 0.117 | 0.147 | 0.109 | 0.124 |
| 2 | 0.203 | 0.181 | 0.173 | 0.176 | 0.180 | 0.168 | 0.183 | 0.197 | 0.190 | 0.190 |
| 3 | 0.245 | 0.236 | 0.241 | 0.230 | 0.235 | 0.226 | 0.244 | 0.245 | 0.264 | 0.251 |
| 4 | 0.287 | 0.290 | 0.306 | 0.282 | 0.289 | 0.280 | 0.299 | 0.292 | 0.333 | 0.308 |
| 5 | 0.326 | 0.342 | 0.367 | 0.334 | 0.342 | 0.331 | 0.350 | 0.337 | 0.396 | 0.361 |
| 6 | 0.365 | 0.391 | 0.425 | 0.385 | 0.393 | 0.378 | 0.395 | 0.382 | 0.454 | 0.410 |
| 7 | 0.402 | 0.439 | 0.479 | 0.435 | 0.443 | 0.421 | 0.436 | 0.425 | 0.505 | 0.455 |
| 8 | 0.438 | 0.485 | 0.530 | 0.485 | 0.492 | 0.461 | 0.471 | 0.468 | 0.551 | 0.497 |
| 9 | 0.472 | 0.529 | 0.577 | 0.533 | 0.539 | 0.497 | 0.501 | 0.509 | 0.591 | 0.534 |
| 10 | 0.505 | 0.570 | 0.620 | 0.581 | 0.585 | 0.529 | 0.526 | 0.549 | 0.625 | 0.567 |
| 11 | 0.537 | 0.610 | 0.660 | 0.628 | 0.629 | 0.558 | 0.546 | 0.588 | 0.653 | 0.596 |
| +gp | 0.615 | 0.705 | 0.746 | 0.756 | 0.746 | 0.667 | 0.616 | 0.652 | 0.717 | 0.662 |

Table 8.3.4 Sole VIIE Stock Weights at Age in kgs

| Age | 1969 | 1970 |
| ---: | :--- | :--- |
| 1 | 0.040 | 0.045 |
| 2 | 0.125 | 0.120 |
| 3 | 0.200 | 0.195 |
| 4 | 0.270 | 0.255 |
| 5 | 0.330 | 0.305 |
| 6 | 0.380 | 0.355 |
| 7 | 0.425 | 0.395 |
| 8 | 0.460 | 0.430 |
| 9 | 0.490 | 0.465 |
| 10 | 0.520 | 0.490 |
| 11 | 0.550 | 0.510 |
| +gp | 0.609 | 0.541 |

Table 8.3.4 Sole VIIE Stock Weights at Age in kgs continued

| Age | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.030 | 0.055 | 0.035 | 0.040 | 0.071 | 0.095 | 0.086 | 0.090 | 0.064 | 0.052 |
| 2 | 0.090 | 0.130 | 0.105 | 0.125 | 0.144 | 0.146 | 0.156 | 0.156 | 0.141 | 0.125 |
| 3 | 0.170 | 0.200 | 0.170 | 0.200 | 0.221 | 0.198 | 0.221 | 0.217 | 0.216 | 0.206 |
| 4 | 0.240 | 0.265 | 0.235 | 0.265 | 0.267 | 0.247 | 0.278 | 0.276 | 0.287 | 0.288 |
| 5 | 0.295 | 0.325 | 0.290 | 0.320 | 0.327 | 0.294 | 0.332 | 0.330 | 0.352 | 0.360 |
| 6 | 0.345 | 0.380 | 0.340 | 0.370 | 0.385 | 0.338 | 0.382 | 0.380 | 0.414 | 0.436 |
| 7 | 0.390 | 0.420 | 0.390 | 0.410 | 0.435 | 0.380 | 0.425 | 0.425 | 0.463 | 0.513 |
| 8 | 0.420 | 0.460 | 0.435 | 0.455 | 0.479 | 0.417 | 0.462 | 0.463 | 0.502 | 0.575 |
| 9 | 0.445 | 0.490 | 0.475 | 0.490 | 0.516 | 0.456 | 0.497 | 0.498 | 0.539 | 0.620 |
| 10 | 0.470 | 0.520 | 0.510 | 0.515 | 0.545 | 0.491 | 0.527 | 0.526 | 0.574 | 0.650 |
| 11 | 0.490 | 0.540 | 0.540 | 0.530 | 0.569 | 0.523 | 0.553 | 0.555 | 0.608 | 0.674 |
| +gp | 0.544 | 0.558 | 0.585 | 0.571 | 0.628 | 0.595 | 0.629 | 0.630 | 0.719 | 0.714 |

Table 8.3.4 Sole VIIE Stock Weights at Age in kgs continued

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.038 | 0.038 | 0.040 | 0.032 | 0.095 | 0.071 | 0.058 | 0.050 | 0.081 | 0.128 |
| 2 | 0.119 | 0.117 | 0.120 | 0.108 | 0.150 | 0.140 | 0.137 | 0.131 | 0.139 | 0.187 |
| 3 | 0.197 | 0.195 | 0.195 | 0.192 | 0.204 | 0.206 | 0.210 | 0.208 | 0.195 | 0.243 |
| 4 | 0.276 | 0.265 | 0.250 | 0.268 | 0.258 | 0.268 | 0.278 | 0.278 | 0.249 | 0.296 |
| 5 | 0.358 | 0.335 | 0.307 | 0.339 | 0.311 | 0.326 | 0.341 | 0.344 | 0.300 | 0.346 |
| 6 | 0.427 | 0.398 | 0.365 | 0.400 | 0.364 | 0.381 | 0.398 | 0.404 | 0.350 | 0.393 |
| 7 | 0.490 | 0.455 | 0.420 | 0.453 | 0.416 | 0.432 | 0.450 | 0.459 | 0.398 | 0.437 |
| 8 | 0.543 | 0.506 | 0.475 | 0.501 | 0.468 | 0.480 | 0.497 | 0.508 | 0.444 | 0.478 |
| 9 | 0.582 | 0.536 | 0.520 | 0.545 | 0.520 | 0.524 | 0.538 | 0.552 | 0.488 | 0.516 |
| 10 | 0.616 | 0.562 | 0.570 | 0.577 | 0.571 | 0.564 | 0.574 | 0.591 | 0.531 | 0.551 |
| 11 | 0.645 | 0.585 | 0.615 | 0.607 | 0.621 | 0.601 | 0.605 | 0.624 | 0.571 | 0.583 |
| +gp | 0.699 | 0.632 | 0.709 | 0.696 | 0.790 | 0.691 | 0.659 | 0.687 | 0.675 | 0.654 |

Table 8.3.4 Sole VIIE Stock Weights at Age in kgs continued

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.065 | 0.048 | 0.000 | 0.091 | 0.103 | 0.139 | 0.000 | 0.000 | 0.132 | 0.110 |
| 2 | 0.144 | 0.128 | 0.114 | 0.153 | 0.163 | 0.189 | 0.156 | 0.162 | 0.183 | 0.172 |
| 3 | 0.219 | 0.202 | 0.178 | 0.212 | 0.221 | 0.238 | 0.215 | 0.220 | 0.233 | 0.230 |
| 4 | 0.290 | 0.272 | 0.239 | 0.268 | 0.275 | 0.285 | 0.272 | 0.274 | 0.280 | 0.284 |
| 5 | 0.355 | 0.336 | 0.296 | 0.322 | 0.326 | 0.331 | 0.327 | 0.325 | 0.326 | 0.333 |
| 6 | 0.416 | 0.395 | 0.350 | 0.372 | 0.374 | 0.376 | 0.380 | 0.374 | 0.369 | 0.379 |
| 7 | 0.473 | 0.449 | 0.401 | 0.419 | 0.419 | 0.420 | 0.431 | 0.419 | 0.410 | 0.421 |
| 8 | 0.524 | 0.498 | 0.448 | 0.463 | 0.461 | 0.463 | 0.480 | 0.462 | 0.448 | 0.458 |
| 9 | 0.572 | 0.542 | 0.492 | 0.505 | 0.500 | 0.504 | 0.526 | 0.501 | 0.485 | 0.492 |
| 10 | 0.614 | 0.580 | 0.532 | 0.543 | 0.536 | 0.544 | 0.570 | 0.537 | 0.519 | 0.521 |
| 11 | 0.652 | 0.613 | 0.570 | 0.578 | 0.568 | 0.583 | 0.612 | 0.571 | 0.551 | 0.546 |
| +gp | 0.731 | 0.677 | 0.659 | 0.659 | 0.641 | 0.677 | 0.717 | 0.650 | 0.624 | 0.643 |

Table 8.3.4 Sole VIIE Stock Weights at Age in kgs continued

| Age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | mean <br> $07-09$ |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.000 | 0.094 | 0.063 | 0.095 | 0.094 | 0.074 | 0.083 | 0.122 | 0.067 | 0.091 |
| 2 | 0.181 | 0.152 | 0.137 | 0.149 | 0.152 | 0.138 | 0.151 | 0.172 | 0.150 | 0.158 |
| 3 | 0.224 | 0.209 | 0.207 | 0.203 | 0.208 | 0.197 | 0.214 | 0.221 | 0.228 | 0.221 |
| 4 | 0.266 | 0.263 | 0.274 | 0.256 | 0.263 | 0.254 | 0.272 | 0.268 | 0.300 | 0.280 |
| 5 | 0.307 | 0.316 | 0.337 | 0.308 | 0.316 | 0.306 | 0.325 | 0.315 | 0.366 | 0.335 |
| 6 | 0.346 | 0.367 | 0.396 | 0.360 | 0.368 | 0.355 | 0.373 | 0.360 | 0.426 | 0.386 |
| 7 | 0.384 | 0.415 | 0.452 | 0.410 | 0.419 | 0.400 | 0.416 | 0.404 | 0.480 | 0.433 |
| 8 | 0.420 | 0.462 | 0.505 | 0.460 | 0.468 | 0.442 | 0.454 | 0.447 | 0.529 | 0.477 |
| 9 | 0.455 | 0.507 | 0.554 | 0.509 | 0.516 | 0.479 | 0.486 | 0.489 | 0.571 | 0.515 |
| 10 | 0.489 | 0.550 | 0.599 | 0.557 | 0.562 | 0.514 | 0.514 | 0.529 | 0.608 | 0.550 |
| 11 | 0.521 | 0.591 | 0.641 | 0.605 | 0.607 | 0.544 | 0.536 | 0.569 | 0.640 | 0.582 |
| +gp | 0.602 | 0.688 | 0.732 | 0.734 | 0.726 | 0.661 | 0.614 | 0.640 | 0.712 | 0.655 |


| $8 \pm 861$ | 0029\％\％ | Lヵて\＆LZ | 8¢792I | 080¢8 2 |  | ${ }^{\text {P7 }}$ OL ${ }_{\text {L }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 69 |  |
| 0 | 0 | 0 | 0 | 0 | 89 |  |
| 0 | 0 | 0 | 0 | 0 | $\angle 9$ |  |
| 0 | 0 | 0 | 0 | 0 | 99 |  |
| 0 | 0 | 0 | 0 | 07 | gs |  |
| 0 | 0 | 0 | 2 | 6 tI |  |  |
| 0 | I69 | 82I | 0 | cs | ¢c |  |
| 0 | 0 | ¢¢L | ¢01 | 79 | \％9 |  |
| 0 | モLI | ¢LI | 881 | 96 | IG |  |
| 0 | 0 | 0 | \％\＆I | 87I | 0 c |  |
| 0 | 28 | 0 | ItI | LIz | $6 \pm$ |  |
| 0 | 28 | 0 | 9\＆\％ | $98 \%$ | 87 |  |
| 0 | 28 | 206 | ¢8I | ¢18 | 27 |  |
| £8 | $99 \%$ | 886 | LIt | 0tit | 97 |  |
| 0 | $0 ¢ t$ | 8LI | 666 | 897I | ¢t |  |
| 2tI | 6801 | 6 ZET | 6SLI | 0298 | 珧 |  |
| 188 | $\ddagger 66$ | 978 | 98.1 | 9Z9t | ¢t |  |
| £86 | 2972 | 9 SO | 99¢T | 2012 | 27 |  |
| でI | cots | 0 0ヶL | 9¢\％\％ | 1976 | It |  |
| coot | ZL62 | 8906 | LLEz | L289I | $0 \pm$ |  |
| 乙19 | 0 tz | 879I | z0t¢ | ¢¢66I | 68 |  |
| むLEL | 21881 | を88ちL | 6187 | 98608 | 88 |  |
| ¢29 | 868tI | L980L | 8962 | ¢8tIt | 28 |  |
| 982 | 0¢cot | Lgiti | ¢012 | ¢9tza | 98 |  |
| £¢01 | ¢8LtI | 98101 | G208 | L08t9 | 98 |  |
| 988 | 690zI | zLcgz | \＆т97I | LG08L | $\pm 8$ |  |
| 268 | ¢28LI | で6Gz | 6L20I | çIE8 | 88 |  |
| Lfet | 9802 | 819LI | ¢6872 | 9788. | 78 |  |
| 9LIz | \％192 | モ¢88L | 69671 | 01892 | 18 |  |
| モ6z\％ | L962 | gLE9I | 289］1 | 99002 | 08 |  |
| L97\％ | ¢998 | 98801 | も¢LtI | 78699 | 67 |  |
| 9¢01 | 78971 | 06 ¢t | GLtEL | 08067 | 87 |  |
| gelt | モ¢Lも | 6789 | c9itl | 9もてもて | 27 |  |
| LZ2 | ZIL9］ | 2 LGOL | ¢888 | 8t97I | 97 |  |
| I¢̧ | 8LZ6I | キ00才 | モも0¢L | 869t | $9 \square$ |  |
| 0 | Z079\％ | 9981 | ¢t9\％ | ${ }^{2} 66$ | ¢ |  |
| 0 | Lut | 7782 | 616 | 188 | ¢ |  |
| 0 | 601 | 028 | 0 | ¢LI | \％ |  |
| 0 | ¢TL | LL6 | 0 | 0 | 17 |  |
| 0 | $07 \pm$ | 0 | 0 | 0 | 07 |  |
| 0 | 60 I | 0 | 0 | 0 | 61 |  |
| Іәчұо | $\mathrm{Imex}^{\text {mex }}$ | ${ }^{\text {s7 }} \mathrm{N}$ | ${ }^{\text {² }}$ | ${ }^{\text {Mre．}} \mathrm{L}^{\text {U }}$ |  |  |

Table 8.3.6 Sole VIIE effort \& CPUE data

| Year | Effort BT u24 | Effort BT o24 | Landings BT u24 | Landings BT o24 | Survey CPUE | $\overline{\mathrm{BTu} 24}$ LPUE | BTo24 <br> LPUE | Survey <br> CPUE <br> MS | BTu24 <br> LPUE <br> MS | BTo24 <br> LPUE <br> MS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 46.33 | 60.90 | 332.79 | 441.99 | 74.24 | 7.18 | 7.26 | 1.28 | 1.49 | 1.96 |
| 1989 | 35.29 | 86.80 | 200.99 | 520.43 | 69.36 | 5.70 | 6.00 | 1.20 | 1.18 | 1.62 |
| 1990 | 36.35 | 78.51 | 238.56 | 474.06 | 43.72 | 6.56 | 6.04 | 0.75 | 1.36 | 1.63 |
| 1991 | 27.93 | 64.94 | 165.12 | 296.01 | 72.58 | 5.91 | 4.56 | 1.25 | 1.22 | 1.23 |
| 1992 | 29.47 | 61.95 | 169.31 | 291.50 | 78.13 | 5.74 | 4.70 | 1.35 | 1.19 | 1.27 |
| 1993 | 31.08 | 65.31 | 199.90 | 281.75 | 49.63 | 6.43 | 4.31 | 0.86 | 1.33 | 1.16 |
| 1994 | 34.77 | 73.47 | 189.29 | 317.87 | 40.66 | 5.44 | 4.33 | 0.70 | 1.13 | 1.17 |
| 1995 | 31.30 | 76.80 | 158.01 | 328.93 | 37.78 | 5.05 | 4.28 | 0.65 | 1.04 | 1.16 |
| 1996 | 33.16 | 94.91 | 164.71 | 300.93 | 48.72 | 4.97 | 3.17 | 0.84 | 1.03 | 0.86 |
| 1997 | 34.15 | 88.68 | 192.26 | 332.09 | 63.11 | 5.63 | 3.74 | 1.09 | 1.17 | 1.01 |
| 1998 | 43.41 | 83.09 | 186.94 | 306.70 | 65.83 | 4.31 | 3.69 | 1.14 | 0.89 | 1.00 |
| 1999 | 42.82 | 73.17 | 185.15 | 271.41 | 54.50 | 4.32 | 3.71 | 0.94 | 0.89 | 1.00 |
| 2000 | 49.07 | 79.58 | 202.29 | 250.02 | 51.94 | 4.12 | 3.14 | 0.90 | 0.85 | 0.85 |
| 2001 | 65.65 | 92.42 | 302.55 | 300.74 | 74.67 | 4.61 | 3.25 | 1.29 | 0.95 | 0.88 |
| 2002 | 61.55 | 92.19 | 293.79 | 298.56 | 43.18 | 4.77 | 3.24 | 0.75 | 0.99 | 0.87 |
| 2003 | 67.25 | 107.01 | 277.64 | 329.50 | 50.28 | 4.13 | 3.08 | 0.87 | 0.85 | 0.83 |
| 2004 | 56.25 | 108.64 | 206.17 | 239.23 | 57.99 | 3.67 | 2.20 | 1.00 | 0.76 | 0.59 |
| 2005 | 51.49 | 107.66 | 198.42 | 255.15 | 35.67 | 3.85 | 2.37 | 0.62 | 0.80 | 0.64 |
| 2006 | 50.87 | 110.87 | 225.31 | 238.63 | 49.10 | 4.43 | 2.15 | 0.85 | 0.92 | 0.58 |
| 2007 | 65.32 | 94.07 | 237.46 | 213.78 | 62.91 | 3.64 | 2.27 | 1.09 | 0.75 | 0.61 |
| 2008 | 76.21 | 83.37 | 222.79 | 170.25 | 73.55 | 2.92 | 2.04 | 1.27 | 0.61 | 0.55 |
| 2009 | 63.66 | 58.99 | 184.35 | 115.31 | 77.38 | 2.90 | 1.95 | 1.34 | 0.60 | 0.53 |

Table 8.3.7: Available tuning information for the assessment process


| 174.7 | 10 | 85 | 158 | 77 | 40 | 2 | 14 | 3 | 6 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 172 | 11 | 104 | 126 | 96 | 49 | 13 | 13 | 12 | 1 |  |  |  |  |
| UK-Inshore |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1973 | 1987 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 15.76 | 28.3 | 142.9 | 145.8 | 28.7 | 28.7 | 33.8 | 4.9 | 15.2 | 8.4 | 1 | 8.4 | 12.7 | 1.2 |
| 12.58 | 17.2 | 117.7 | 67.5 | 51.6 | 18 | 19.3 | 11 | 8.2 | 5.8 | 12 | 3.1 | 4.8 | 2.9 |
| 12.84 | 30 | 163.3 | 41.9 | 45.1 | 21.2 | 4.8 | 10 | 4.9 | 3.7 | 3.7 | 7 | 3.8 | 5.2 |
| 12.58 | 63.6 | 137.5 | 139.9 | 44.9 | 32.6 | 21.4 | 11.4 | 14.4 | 11.7 | 2.9 | 3.7 | 16 | 4.6 |
| 14.01 | 169.7 | 106.7 | 114.5 | 57.4 | 24.3 | 15.8 | 18.1 | 2.5 | 5.3 | 6.4 | 3.5 | 4.5 | 8.2 |
| 22.31 | 117.8 | 449.7 | 124.4 | 72.1 | 54.5 | 28.5 | 21.1 | 22.5 | 10.4 | 6.7 | 5.8 | 5.9 | 3.5 |
| 31.15 | 114.2 | 342.9 | 310.5 | 89.6 | 70.2 | 51.1 | 32.4 | 28.1 | 30.2 | 7.3 | 6.8 | 17.3 | 3.6 |
| 42.4 | 131.4 | 322.7 | 221.1 | 257.7 | 36.9 | 46.3 | 37.1 | 18.1 | 13.7 | 32.5 | 9.2 | 7.6 | 8.9 |
| 46.36 | 161.9 | 478.9 | 320.6 | 190.5 | 123.1 | 52.6 | 37.8 | 22.1 | 15.7 | 12.1 | 11.3 | 3.4 | 3.7 |
| 51.68 | 86 | 857.6 | 442 | 215.7 | 113.5 | 70.6 | 43 | 33.6 | 22.2 | 16.7 | 10.3 | 8.2 | 7.6 |
| 51.09 | 76.8 | 353.4 | 623.5 | 210.6 | 80.1 | 78.3 | 94.1 | 33.8 | 26.4 | 5.3 | 6.5 | 34.8 | 5.1 |
| 48.21 | 177.7 | 280.2 | 309 | 257 | 88.6 | 43.9 | 39.6 | 38.1 | 8.5 | 5.9 | 13.9 | 17.5 | 4 |
| 54.87 | 57.7 | 598.4 | 320.7 | 168.7 | 198.1 | 37.2 | 29.9 | 45.9 | 32.4 | 17.7 | 7.6 | 4.2 | 5.6 |
| 53.46 | 103.2 | 823.1 | 361.7 | 111.3 | 82.9 | 87.1 | 23.2 | 9.3 | 7.6 | 17.8 | 4.2 | 5.1 | 9.4 |
| 35.61 | 116.6 | 183.2 | 269.3 | 93.4 | 17.1 | 16.7 | 32 | 5.9 | 9 | 3.6 | 7.8 | 4.5 | 5.2 |
| UK-Offshore |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1973 | 1987 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 3 | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| 5.64 | 24.6 | 37.3 | 8.9 | 13 | 16.8 | 2.1 | 6.6 | 4.3 | 0.7 | 4.3 | 4.3 | 0.7 |  |
| 6.72 | 30.3 | 25.7 | 23.8 | 12.2 | 14.4 | 7.1 | 5.4 | 4.5 | 11.3 | 2.3 | 2.4 | 2.4 |  |
| 13.94 | 85.2 | 32.5 | 42.1 | 29.2 | 7.3 | 13.1 | 6.4 | 5.8 | 6.9 | 10.8 | 3.8 | 8.7 |  |
| 7.36 | 38.6 | 58.4 | 22.7 | 24.2 | 17.3 | 8.1 | 10.2 | 9.8 | 2.9 | 3 | 8.8 | 4.2 |  |
| 9.88 | 36.1 | 57.7 | 34.9 | 21.7 | 15.5 | 15.3 | 2.1 | 5.3 | 7.9 | 3.5 | 3 | 8.8 |  |
| 14.5 | 140.5 | 57.7 | 40.4 | 44.9 | 25.8 | 16.6 | 17.9 | 9.7 | 7.7 | 5.3 | 3.6 | 3.5 |  |
| 20.38 | 107.9 | 145.1 | 50.6 | 58.2 | 46.4 | 25.5 | 22.4 | 28.3 | 8.3 | 6.3 | 10.6 | 3.7 |  |
| 28.18 | 103.1 | 104.9 | 147.7 | 31.1 | 42.7 | 29.7 | 14.7 | 13 | 37.9 | 8.8 | 4.7 | 9 |  |
| 28.75 | 142.8 | 142.1 | 101.9 | 96.6 | 45.3 | 28.2 | 16.7 | 13.9 | 13.1 | 10 | 2 | 3.5 |  |
| 39.85 | 317.9 | 243.4 | 143.3 | 110.7 | 75.7 | 39.9 | 31.6 | 24.5 | 22.5 | 11.3 | 5.9 | 9 |  |
| 66.45 | 104.1 | 433.6 | 167.6 | 116.5 | 100.9 | 104.4 | 47.8 | 27.7 | 19.8 | 9.2 | 18.7 | 10.2 |  |
| 49.07 | 152.8 | 234.7 | 214.8 | 133.2 | 69.9 | 22.9 | 54.3 | 28.5 | 7.8 | 29.7 | 8.2 | 6.7 |  |
| 47.15 | 245.2 | 130.3 | 110.8 | 211.1 | 75.6 | 26.7 | 31.6 | 15.5 | 7.1 | 0 | 7.9 | 6.8 |  |
| 34.66 | 425.5 | 215.7 | 100.2 | 79.1 | 70 | 15.2 | 7.9 | 30.1 | 28.6 | 5.3 | 13.7 | 7.6 |  |
| 47.41 | 158.4 | 344.2 | 138.8 | 53.3 | 50.7 | 95.7 | 22.7 | 19 | 26.1 | 13.8 | 14.2 | 14.6 |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics FLR XSA Diagnostics 2010-05-20 17:50:34

CPUE data from index.final
Catch data for 41 years. 1969 to 2009. Ages 1 to 12.

| fleet | first last first last alpha beta age age year year |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK-CBT | 3 | 11 | 1988 | 2009 | 0 | 1 |
| UK-COT | 3 | 11 | 1988 | 2009 | 0 | 1 |
| UK-WEC-BTS | 1 | 9 | 1988 | 2009 | 0.75 | 0.8 |
| UK-Inshore | 2 | 11 | 1973 | 1987 | 0 | 1 |
| UK-Offshore | 3 | 11 | 1973 | 1987 | 0 | 1 |

Time series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages $>7$

Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 10 years or
the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=0.5$
min. S.E. for population estimates derived from each fleet $=0.5$
Regression weights
Year

| 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Estimated population abundance at 1st Jan 2010
Age

| 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 4753 | 2371 | 1765 | 1428 | 1018 | 510 | 190 | 263 | 86 | 128 | 110 |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued
XSA fleet diagnostics for UK-CBT


| 30.299 | 0.012 | 0.222 | 0.052 | 20.077 | -0.069 | 0.103 | -0.373 | -0.011 | 0.34 |  | 0.004 | 0.105 | -0.340 | 0.027 | -0.271 | $1-0.505$ | -0.599 | 0.367 | 0.013 | 0.176 | -0.032 | 0.401 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $4 \quad 0.322$ | 0.225 | 50.215 | -0.148 | -0.172 | -0.001 | 0.186 | 0.272 | 20.230 | 0.12 |  | 0.062 | -0.014 | 0.011 | -0.294 | -0.631 | -0.301 | $-0.767$ | 0.395 | 0.058 | 0.090 | -0.004 | 0.146 |
| $5 \quad 0.367$ | 0.290 | 0.189 | -0.205 | -0.364 | 0.206 | 0.118 | 0.107 | 70.094 | 0.28 | - | -0.028 | 0.168 | -0.050 | -0.006 | -0.506 | -0.345 | -0.790 | 0.209 | 0.091 | 0.209 | -0.195 | 0.153 |
| 60.413 | 0.357 | 0.460 | -0.363 | -1.043 | 0.049 | -0.054 | -0.106 | 60.015 | -0.008 | 08 | 0.272 | 0.147 | -0.110 | 0.064 | 0.065 | -0.812 | -0.237 | 0.503 | -0.171 | 0.044 | 0.241 | 0.272 |
| 7 -0.091 | 0.386 | 0.161 | -0.644 | -0.633 | 0.127 | 0.304 | 0.200 | -0.067 | 0.36 |  | -0.040 | 0.420 | -0.334 | 0.140 | 0.010 | -0.449 | -1.197 | 0.603 | 0.168 | 0.291 | 0.126 | 0.155 |
| 80.074 | 0.310 | 0.139 | -0.539 | -0.929 | -0.137 | -0.136 | 0.259 | 90.125 | -0.10 | - | -0.228 | 0.502 | -0.106 | -0.029 | -0.121 | -0.430 | -0.429 | -0.035 | 0.473 | 0.501 | 0.651 | 0.185 |
| $9-0.281$ | 0.433 | 0.547 | -0.766 | -1.032 | 0.148 | 0.309 | -0.082 | 20.133 | -0.039 |  | -0.240 | -0.059 | 0.217 | 0.497 | -0.666 | -0.026 | -0.592 | 0.679 | 0.216 | 0.439 | 0.423 | 0.491 |
| $10-0.972$ | 0.171 | 0.177 | -0.207 | 7 -1.298 | -0.422 | 0.634 | -0.273 | 30.104 | -0.61 |  | -0.272 | 0.204 | -0.028 | 0.206 | -0.104 | -0.211 | -0.374 | 0.703 | 0.118 | 0.144 | 0.549 | 0.395 |
| 110.173 | 0.783 | -0.138 | -0.678 | - -1.299 | 0.220 | 0.160 | -0.351 | 10.055 | -0.042 |  | 0.245 | 0.382 | 0.045 | 0.159 | -0.044 | -0.020 | 0.027 | 0.906 | 0.128 | 0.222 | 0.721 | 0.431 |
| Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Age3 |  | Age 4 |  | dge5 |  | Age6 |  | Age' |  | Age8 |  | Age9 |  | Age10 |  | Age11 |  |  |  |  |
| MeanLogq |  | 15.927 |  | 5.8395 | -15.8 | 8656 | -15.91 | 9101 | -16.0 | 0115 |  | -16.0766 |  | -16.0766 |  | -16.0766 |  | 16.0766 |  |  |  |  |
| S.ELogq |  | 0.2726 |  | 0.2915 |  | 2909 |  | 3775 |  | 4223 |  | 0.380 |  | 0.4684 |  | 0.4897 |  | 0.4701 |  |  |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued
XSA fleet diagnostics for UK-WEC-BTS

\footnotetext{


Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## XSA fleet diagnostics for UK-Inshore

| Age | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.186 | -0.408 | 0.200 | 1.055 | 1.102 | 0.617 | 0.325 | 0.057 | -0.377 | -0.569 | -0.459 | -0.002 | -1.439 | -0.203 | -0.085 |
| 3 | 0.361 | 0.483 | 0.477 | 0.380 | 0.105 | 0.300 | 0.034 | -0.295 | -0.070 | -0.149 | -0.490 | -0.444 | -0.145 | -0.043 | -0.506 |
| 4 | 0.271 | 0.098 | -0.324 | 0.680 | 0.376 | 0.060 | -0.074 | -0.401 | -0.077 | 0.139 | -0.056 | -0.206 | -0.038 | -0.108 | -0.338 |
| 5 | 0.145 | -0.029 | 0.193 | 0.258 | 0.261 | 0.003 | -0.030 | 0.049 | -0.040 | 0.100 | 0.182 | -0.172 | -0.262 | -0.322 | -0.335 |
| 6 | 0.231 | 0.482 | -0.410 | 0.486 | 0.104 | 0.360 | 0.280 | -0.613 | -0.086 | 0.057 | -0.072 | 0.223 | 0.200 | -0.213 | -1.029 |
| 7 | 0.561 | 0.414 | -0.521 | -0.067 | 0.013 | 0.163 | 0.419 | 0.045 | 0.087 | -0.293 | 0.258 | -0.021 | -0.164 | -0.026 | -0.868 |
| 8 | -0.423 | -0.156 | -0.083 | 0.538 | -0.196 | 0.016 | 0.198 | 0.133 | 0.099 | 0.132 | 0.359 | 0.015 | -0.166 | -0.238 | -0.228 |
| 9 | 0.369 | 0.511 | -0.826 | 0.523 | -0.895 | -0.257 | 0.233 | -0.410 | -0.106 | 0.300 | 0.361 | -0.048 | 0.415 | -0.947 | -0.878 |
| 10 | -0.571 | -0.169 | -0.081 | 0.245 | -0.378 | 0.270 | -0.068 | -0.505 | -0.363 | 0.187 | 0.565 | -0.493 | 0.033 | -0.899 | -0.205 |
| 11 | -0.180 | 0.193 | -0.459 | -0.103 | -0.264 | -0.426 | -0.146 | 0.017 | -0.422 | -0.072 | -0.816 | -0.446 | 0.484 | -0.167 | -0.771 |

Mean log catchability and standard error of ages with catchability

|  | Age2 | Age3 | Age4 | Age5 | Age6 | Age7 | Age8 | Age9 | Age10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| MeanLogq | -7.3059 | -5.7416 | -5.7272 | -5.9856 | -6.3199 | -6.4512 | -6.4189 | -6.4189 | -6.4189 |
| S.ELogq | 0.6466 | 0.3452 | 0.2878 | 0.1993 | 0.4221 | 0.3692 | 0.2518 | 0.5564 | 0.3857 |
|  |  |  |  |  | 0.3189 |  |  |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued
XSA fleet diagnostics for UK-Offshore


Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued
Year Class 2008 at terminal Age 1


Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2007 at terminal Age 2



Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2006 at terminal Age 3



Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2005 at terminal Age 4



Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2004 at terminal Age 5

| Source fshk | Age 1 | $\text { Age } 2$ | $\text { Age } 3$ | Age 4 <br> 1 | Age 5 $616$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 4.0000 |  |  |  |  |
| UK-CBT | 1 | 1 | 949 | 764 | 745 |  |  |  |  |
|  | 0.0000 | 0.0000 | 1.5952 | 2.1555 | 3.1184 |  |  |  |  |
| UK-COT | 1 | 1 | 1214 | 1014 | 1186 |  |  |  |  |
|  | 0.0000 | 0.0000 | 1.5952 | 2.1555 | 3.1184 |  |  |  |  |
| $\begin{aligned} & \text { UK- } \\ & \text { WEC- } \\ & \text { BTS } \end{aligned}$ | 748 | 1130 | 1257 | 1540 | 1861 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
|  | 0.9701 | 1.4016 | 1.5952 | 2.1555 | 3.1184 |  |  |  |  |
| Source |  | Survi | ivors in | int s.e. e | ext s.e. | Var <br> Ratio | N | Scaled W | F est. |
| fshk |  |  | 616 | 0.441 | NaN | NaN | 1 | 0.148 | 0.382 |
| UK-CBT |  |  | 795 | 0.299 | 0.069 | 0.232 | 3 | 0.255 | 0.308 |
| UK-COT |  |  | 1135 | 0.299 | 0.055 | 0.182 | 3 | 0.255 | 0.225 |
| UK-WEC-BTS |  |  | 1402 | 0.243 | 0.141 | 0.579 | 5 | 0.343 | 0.186 |
| term. Surv. |  | int s.e | e. ext s | s.e. | N Var | Ratio | F |  |  |
| 1018 |  | 0.151 | 10.1 | 131 | 12 Va | r Ratio | 0.248 |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2003 at terminal Age 6



Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2002 at terminal Age 7

| Source fshk | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 1 | 1 | 1 | 1 | 1 | 144 |  |  |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 4.0000 |  |  |
| UK-CBT | 1 | 1 | 134 | 145 | 152 | 180 | 147 |  |  |
|  | 0.0000 | 0.0000 | 0.6096 | 0.8014 | 1.2332 | 1.9552 | 3.0548 |  |  |
| UK-COT | 1 | 1 | 275 | 202 | 234 | 242 | 222 |  |  |
|  | 0.0000 | 0.0000 | 0.6096 | 0.8014 | 1.2332 | 1.95523 | 3.0548 |  |  |
| UK- <br> WECBTS | 708 | 360 | 108 | 157 | 151 | 71 | 697 |  |  |
|  |  |  |  |  |  |  |  |  |  |
|  | 0.3564 | 0.5150 | 0.6096 | 0.8014 | 1.2332 | 1.5636 | 2.1426 |  |  |
| Source |  | Survi | ivors i | int s.e. e | ext s.e. | Var <br> Ratio | N | Scaled W | F est. |
| fshk |  |  | 144 | 0.437 | NaN | NaN | 1 | 0.151 | 0.343 |
| UK-CBT |  |  | 154 | 0.259 | 0.048 | 0.184 | 5 | 0.289 | 0.322 |
| UK-COT |  |  | 231 | 0.259 | 0.037 | 0.144 | 5 | 0.289 | 0.227 |
| UK-WEC-BTS |  |  | 226 | 0.246 | 0.373 | 1.518 | 7 | 0.272 | 0.231 |
| term. Surv. |  | int s.e | e. ext s |  | N Var | r. Ratio | F |  |  |
| 190 |  | 0.141 | 10.1 | 107 | 18 Var | ar Ratio | 0.269 |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2001 at terminal Age 8

| Source fshk | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 |  | $\text { Age } 8$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  | 156 |  |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |  | 4.0000 |  |
| UK-CBT | 1 | 1 | 193 | 182 | 208 | 176 | 208 |  | 177 |  |
|  | 0.0000 | 0.0000 | 0.4475 | 0.6090 | 0.8805 | 1.3710 | 2.1684 |  | 3.2203 |  |
| UK-COT | 1 | 1 | 144 | 390 | 288 | 275 | 298 |  | 316 |  |
|  | 0.0000 | 0.0000 | 0.4475 | 0.6090 | 0.8805 | 1.3710 | 2.1684 |  | 3.2203 |  |
| UK- <br> WECBTS | 121 | 335 | 256 | 208 | 318 | 505 | 523 |  | 728 |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | 0.2734 | 0.3950 | 0.4475 | 0.6090 | 0.8805 | 1.0964 | 1.5209 |  | 2.2944 |  |
| Source |  | Surv | vivors in | nt s.e. e | ext s.e. | Var <br> Ratio |  |  | Scaled W | F est. |
| fshk |  |  | 156 | 0.449 | NaN | NaN |  | 1 | 0.138 | 0.341 |
| UK-CBT |  |  | 188 | 0.246 | 0.034 | 0.138 |  | 6 | 0.301 | 0.290 |
| UK-COT |  |  | 294 | 0.246 | 0.083 | 0.337 |  | 6 | 0.301 | 0.195 |
| UK-WEC-BTS |  |  | 447 | 0.247 | 0.179 | 0.726 |  | 8 | 0.260 | 0.133 |
| term. Surv. |  | int s.e | e. ext s |  | N Var. | . Ratio |  | F |  |  |
| 263 |  | 0.13 | 380.1 | 13 | 21 Var | ar Ratio | 0.21 |  |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 2000 at terminal Age 9



Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 1999 at terminal Age 10

| Source | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 |  | Age 8 | Age 9 | Age 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| fshk | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 93 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |  | 0.0000 | 0.0000 | 4.0000 |
| UK-CBT | 1 | 1 | 149 | 125 | 68 | 107 | 125 |  | 136 | 133 | 98 |
|  | 0.0000 | 0.0000 | 0.2475 | 0.3464 | 0.5220 | 0.7067 | 1.0417 |  | 1.5325 | 2.1973 | 2.9823 |
| UK-COT | 1 | 1 | 98 | 95 | 58 | 212 | 151 |  | 211 | 196 | 190 |
|  | 0.0000 | 0.0000 | 0.2475 | 0.3464 | 0.5220 | 0.7067 | 1.0417 |  | 1.5325 | 2.1973 | 2.9249 |
| UK- <br> WECBTS | 121 | 185 | 102 | 103 | 63 | 102 | 97 | 7 | 75 | 237 | 1 |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.1663 | 0.2403 | 0.2475 | 0.3464 | 0.5220 | 0.5651 | 0.7306 |  | 1.0919 | 1.4243 | 0.0000 |
| Source |  | Surv | vivors in | int s.e. e | ext s.e. | Var Ratio |  |  | Scal | ed W | $F$ est. |
| fshk |  |  | 93 | 0.432 | NaN | NaN |  | 1 |  | 0.141 | 0.383 |
| UK-CBT |  |  | 115 | 0.222 | 0.071 | 0.319 |  | 8 |  | 0.337 | 0.322 |
| UK-COT |  |  | 172 | 0.222 | 0.122 | 0.549 |  | 8 |  | 0.335 | 0.226 |
| UK-WEC-BTS |  |  | 117 | 0.239 | 0.169 | 0.707 |  | 9 |  | 0.188 | 0.316 |
| term. Surv. |  | int s.e | e. ext s |  | N Var. | r. Ratio |  | F |  |  |  |
| 128 |  | 0.1 | 30. | . 07 | 26 Var | ar Ratio | 0.29 | 293 |  |  |  |

Table 3.2.8 Sole VIIE XSA detailed survivor diagnostics continued

## Year Class 1998 at terminal Age 11

| Source | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Age 8 | Age 9 Age 10 Age 11 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| fshk | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4.0000 |  |  |  |  |  |  |  |  |  |  |
| UK-CBT | 1 | 1 | 111 | 80 | 94 | 100 | 103 | 137 | 114 | 126 |
|  | 0.0000 | 0.0000 | 0.1966 | 0.2685 | 0.3546 | 0.4866 | 0.7323 | 1.0627 | 1.5503 | 2.2198 |
| 3.1389 |  |  |  |  |  |  |  |  |  |  |
| UK-COT | 1 | 1 | 113 | 59 | 78 | 87 | 201 | 177 | 171 | 191 |
|  | 0.0000 | 0.0000 | 0.1966 | 0.2685 | 0.3546 | 0.4866 | 0.7323 | 1.0627 | 1.5503 | 2.1770 |
| 3.1389 |  |  |  |  |  |  |  |  |  |  |
| UK- | 75 | 90 | 105 | 59 | 48 | 61 | 52 | 141 | 25 | 1 |

WEC-
BTS
$\begin{array}{lllllllllll}0.1292 & 0.1867 & 0.1966 & 0.2685 & 0.3546 & 0.3892 & 0.5137 & 0.7571 & 1.0049 & 0.0000 & 0.0000\end{array}$

| Source | Survivors int s.e. ext s.e. | Var <br> Ratio | N | Scaled W | F est. |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| fshk | 93 | 0.443 | NaN | NaN | 1 | 0.144 | 0.281 |
| UK-CBT | 103 | 0.220 | 0.071 | 0.323 | 9 | 0.360 | 0.256 |
| UK-COT | 161 | 0.220 | 0.100 | 0.453 | 9 | 0.359 | 0.172 |
| UK-WEC-BTS | 58 | 0.236 | 0.216 | 0.914 | 9 | 0.137 | 0.420 |
|  |  |  |  |  |  |  |  |
| term. Surv. | int s.e. ext s.e. | N Var. Ratio | F |  |  |  |  |
| 110 | 0.133 | 0.063 | 28 | Var Ratio | 0.242 |  |  |

Table 8.3.9 Sole VIIE Stock Numbers at Age in 000's

| Age | 1969 | 1970 | 1971 | 1972 |
| ---: | ---: | ---: | ---: | ---: |
| 1 | 1609 | 3974 | 2954 | 2618 |
| 2 | 2179 | 1456 | 3596 | 2673 |
| 3 | 2437 | 1887 | 1267 | 3206 |
| 4 | 761 | 1899 | 1487 | 955 |
| 5 | 1072 | 613 | 1412 | 1112 |
| 6 | 1671 | 829 | 469 | 1089 |
| 7 | 181 | 1312 | 671 | 363 |
| 8 | 583 | 143 | 1081 | 531 |
| 9 | 667 | 480 | 117 | 830 |
| 10 | 298 | 579 | 402 | 97 |
| 11 | 102 | 250 | 475 | 331 |
| gp | 720 | 1291 | 981 | 653 |
| Total | 12279 | 14714 | 14912 | 14457 |

Table 8.3.9 Sole VIIE Stock Numbers at Age in 000's continued

| Age | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 3580 | 3357 | 3142 | 7205 | 5071 | 4713 | 5163 | 8942 | 5157 | 4164 |
| 2 | 2369 | 3239 | 3037 | 2843 | 6519 | 4589 | 4265 | 4672 | 8091 | 4666 |
| 3 | 2280 | 2076 | 2888 | 2670 | 2413 | 5494 | 3914 | 3643 | 4060 | 7089 |
| 4 | 2509 | 1686 | 1546 | 2074 | 2017 | 1881 | 3902 | 2778 | 2764 | 2907 |
| 5 | 705 | 1858 | 1317 | 1237 | 1428 | 1460 | 1372 | 2759 | 2041 | 1882 |
| 6 | 896 | 554 | 1512 | 1002 | 966 | 1096 | 1118 | 1004 | 1898 | 1402 |
| 7 | 878 | 717 | 433 | 1259 | 778 | 777 | 812 | 794 | 789 | 1347 |
| 8 | 315 | 681 | 573 | 366 | 1052 | 638 | 605 | 570 | 544 | 544 |
| 9 | 457 | 269 | 575 | 468 | 287 | 881 | 509 | 449 | 382 | 372 |
| 10 | 624 | 364 | 213 | 496 | 368 | 250 | 724 | 375 | 345 | 274 |
| 11 | 51 | 536 | 307 | 172 | 400 | 311 | 190 | 556 | 286 | 257 |
| + gp | 1695 | 1034 | 2169 | 2629 | 1993 | 2309 | 1972 | 1456 | 1100 | 1241 |
| Total | 16360 | 16371 | 17713 | 22422 | 23293 | 24398 | 24545 | 27998 | 27457 | 26145 |

Table 8.3.9 Sole VIIE Stock Numbers at Age in 000's continued

| Age | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 6580 | 7836 | 4233 | 6446 | 4203 | 4114 | 3146 | 7902 | 4367 |
| 2 | 3768 | 5954 | 7090 | 3831 | 5832 | 3803 | 3722 | 2846 | 7150 |
| 3 | 4101 | 3323 | 5070 | 6142 | 3232 | 4814 | 3020 | 2997 | 2251 |
| 4 | 5034 | 2994 | 2376 | 2970 | 4019 | 2155 | 2988 | 1904 | 1854 |
| 5 | 1759 | 3058 | 1924 | 1431 | 1764 | 2599 | 1384 | 1531 | 1169 |
| 1218 |  |  |  |  |  |  |  |  |  |
| 6 | 1176 | 1037 | 2047 | 1294 | 895 | 1190 | 1659 | 779 | 860 |
| 7833 |  |  |  |  |  |  |  |  |  |
| 7 | 934 | 730 | 630 | 1296 | 866 | 616 | 721 | 1017 | 473 |
| 8 | 991 | 591 | 467 | 400 | 853 | 571 | 411 | 438 | 669 |
| 9 | 363 | 617 | 412 | 330 | 282 | 554 | 362 | 268 | 260 |
| 10 | 230 | 216 | 414 | 275 | 227 | 210 | 397 | 226 | 145 |
| 11 | 171 | 138 | 144 | 294 | 163 | 158 | 153 | 252 | 134 |
| gp | 1210 | 1239 | 623 | 503 | 644 | 539 | 687 | 695 | 630 |
| +otal | 26316 | 27735 | 25429 | 25212 | 22980 | 21322 | 18651 | 20856 | 19962 |

Table 8.3.9 Sole VIIE Stock Numbers at Age in 000's continued

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 2596 | 3721 | 4384 | 3663 | 4851 | 3846 | 7079 | 5796 | 4090 | 5922 |
| 2 | 3441 | 2349 | 3367 | 3967 | 3314 | 4389 | 3480 | 6405 | 5245 | 3700 |
| 3 | 3276 | 2915 | 2033 | 2956 | 3243 | 2793 | 3719 | 2883 | 5503 | 4608 |
| 4 | 4103 | 2295 | 2013 | 1547 | 2252 | 2143 | 1951 | 2496 | 2039 | 3647 |
| 5 | 1089 | 2659 | 1545 | 1223 | 1053 | 1349 | 1429 | 1290 | 1544 | 1339 |
| 6 | 820 | 652 | 1875 | 992 | 823 | 644 | 901 | 915 | 818 | 924 |
| 7 | 599 | 533 | 484 | 1306 | 674 | 574 | 383 | 572 | 610 | 486 |
| 8 | 407 | 399 | 382 | 313 | 937 | 425 | 375 | 238 | 416 | 383 |
| 9 | 249 | 294 | 314 | 250 | 219 | 683 | 309 | 242 | 165 | 281 |
| 10 | 382 | 169 | 212 | 226 | 168 | 156 | 497 | 228 | 155 | 97 |
| 11 | 130 | 293 | 111 | 160 | 158 | 133 | 108 | 348 | 158 | 99 |
| + gp | 330 | 574 | 850 | 625 | 345 | 623 | 536 | 361 | 592 | 459 |
| Total | 17424 | 16852 | 17569 | 17228 | 18036 | 17760 | 20769 | 21774 | 21334 | 21945 |

Table 8.3.9 Sole VIIE Stock Numbers at Age in 000's continued

| Age | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 <br> sur- <br> vivors | geom <br> mean <br> $03-09$ | arith <br> mean <br> $03-09$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 2976 | 4225 | 4789 | 3811 | 3214 | 3038 | $4332^{a}$ | 0 | 3815 | 3901 |
| 2 | 5359 | 2693 | 3823 | 4333 | 3448 | 2908 | 2749 | 3920 | 3512 | 3616 |
| 3 | 3032 | 4280 | 2058 | 3214 | 3445 | 2929 | 2364 | 2371 | 2972 | 3046 |
| 4 | 2979 | 1949 | 2845 | 1417 | 2160 | 2307 | 1935 | 1765 | 2169 | 2227 |
| 5 | 2499 | 1789 | 1338 | 1781 | 833 | 1236 | 1443 | 1428 | 1484 | 1560 |
| 6 | 844 | 1647 | 1196 | 783 | 1035 | 475 | 761 | 1018 | 902 | 963 |
| 7 | 530 | 640 | 990 | 734 | 470 | 592 | 275 | 510 | 567 | 604 |
| 8 | 317 | 409 | 422 | 618 | 451 | 255 | 361 | 190 | 391 | 405 |
| 9 | 246 | 240 | 259 | 275 | 383 | 285 | 131 | 263 | 249 | 260 |
| 10 | 165 | 168 | 159 | 159 | 167 | 242 | 190 | 86 | 177 | 178 |
| 11 | 57 | 89 | 100 | 101 | 91 | 99 | 155 | 128 | 95 | 99 |
| + gp | 252 | 214 | 395 | 283 | 335 | 223 | 286 | 313 | 278 | 284 |
| Total | 19254 | 18341 | 18375 | 17508 | 16033 | 14590 | 15903 |  |  |  |

[^19]Table 8.3.10 Sole VIIE Fishing Mortality at Age

| Age | 1969 | 1970 |  |
| ---: | :---: | :---: | :---: |
| 1 | 0.000 | 0.000 |  |
| 2 |  | 0.044 | 0.039 |
| 3 |  | 0.149 | 0.138 |
| 4 |  | 0.117 | 0.196 |
| 5 | 0.157 | 0.167 |  |
| 6 |  | 0.141 | 0.111 |
| 7 | 0.132 | 0.094 |  |
| 8 |  | 0.094 | 0.099 |
| 9 | 0.041 | 0.079 |  |
| 10 |  | 0.074 | 0.099 |
| 11 | 0.096 | 0.097 |  |
| + gp | 0.096 | 0.097 |  |
| Fbar3-9 | 0.119 | 0.126 |  |

Table 8.3.10 Sole VIIE Fishing Mortality at Age continued

| Age | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 0.015 | 0.059 | 0.032 | 0.015 | 0.029 | 0.064 | 0.071 | 0.059 | 0.058 | 0.040 |
| 3 | 0.182 | 0.145 | 0.202 | 0.195 | 0.231 | 0.180 | 0.149 | 0.242 | 0.243 | 0.176 |
| 4 | 0.191 | 0.203 | 0.200 | 0.148 | 0.123 | 0.274 | 0.223 | 0.215 | 0.247 | 0.208 |
| 5 | 0.160 | 0.116 | 0.142 | 0.106 | 0.173 | 0.147 | 0.164 | 0.167 | 0.212 | 0.274 |
| 6 | 0.158 | 0.115 | 0.123 | 0.145 | 0.083 | 0.153 | 0.118 | 0.200 | 0.242 | 0.141 |
| 7 | 0.134 | 0.042 | 0.155 | 0.124 | 0.069 | 0.080 | 0.099 | 0.151 | 0.254 | 0.277 |
| 8 | 0.164 | 0.050 | 0.057 | 0.068 | 0.102 | 0.143 | 0.077 | 0.126 | 0.197 | 0.299 |
| 9 | 0.092 | 0.185 | 0.126 | 0.134 | 0.048 | 0.141 | 0.038 | 0.096 | 0.206 | 0.164 |
| 10 | 0.095 | 0.540 | 0.052 | 0.073 | 0.114 | 0.115 | 0.070 | 0.175 | 0.164 | 0.171 |
| 11 | 0.129 | 0.187 | 0.088 | 0.114 | 0.086 | 0.089 | 0.087 | 0.097 | 0.167 | 0.282 |
| + gp | 0.129 | 0.187 | 0.088 | 0.114 | 0.086 | 0.089 | 0.087 | 0.097 | 0.167 | 0.282 |
| Fbar3-9 | 0.154 | 0.122 | 0.144 | 0.131 | 0.119 | 0.160 | 0.124 | 0.171 | 0.229 | 0.220 |

Table 8.3.10 Sole VIIE Fishing Mortality at Age continued

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| 1 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 0.032 | 0.029 | 0.026 | 0.061 | 0.043 | 0.070 | 0.092 | 0.131 | 0.117 | 0.135 |
| 3 | 0.234 | 0.242 | 0.214 | 0.235 | 0.435 | 0.324 | 0.305 | 0.377 | 0.361 | 0.380 |
| 4 | 0.285 | 0.403 | 0.398 | 0.343 | 0.407 | 0.421 | 0.336 | 0.343 | 0.569 | 0.387 |
| 5 | 0.275 | 0.370 | 0.428 | 0.302 | 0.296 | 0.370 | 0.294 | 0.349 | 0.474 | 0.477 |
| 6 | 0.243 | 0.306 | 0.376 | 0.399 | 0.357 | 0.302 | 0.274 | 0.401 | 0.389 | 0.399 |
| 7 | 0.272 | 0.207 | 0.358 | 0.347 | 0.354 | 0.318 | 0.317 | 0.303 | 0.398 | 0.319 |
| 8 | 0.280 | 0.305 | 0.374 | 0.261 | 0.248 | 0.249 | 0.332 | 0.354 | 0.330 | 0.419 |
| 9 | 0.233 | 0.384 | 0.421 | 0.300 | 0.304 | 0.275 | 0.193 | 0.232 | 0.371 | 0.516 |
| 10 | 0.195 | 0.370 | 0.406 | 0.306 | 0.242 | 0.422 | 0.263 | 0.216 | 0.355 | 0.427 |
| 11 | 0.225 | 0.285 | 0.257 | 0.243 | 0.237 | 0.307 | 0.307 | 0.403 | 0.395 | 0.436 |
| + gp | 0.225 | 0.285 | 0.257 | 0.243 | 0.237 | 0.307 | 0.307 | 0.403 | 0.395 | 0.436 |
| Fbar3-9 | 0.260 | 0.317 | 0.367 | 0.312 | 0.343 | 0.323 | 0.293 | 0.337 | 0.413 | 0.414 |

Table 8.3.10 Sole VIIE Fishing Mortality at Age continued

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 0.068 | 0.088 | 0.066 | 0.045 | 0.030 | 0.102 | 0.071 | 0.066 | 0.088 | 0.052 |
| 3 | 0.215 | 0.287 | 0.256 | 0.270 | 0.174 | 0.172 | 0.314 | 0.259 | 0.299 | 0.246 |
| 4 | 0.320 | 0.310 | 0.334 | 0.295 | 0.399 | 0.284 | 0.412 | 0.305 | 0.314 | 0.380 |
| 5 | 0.301 | 0.296 | 0.413 | 0.249 | 0.343 | 0.296 | 0.392 | 0.304 | 0.346 | 0.355 |
| 6 | 0.313 | 0.167 | 0.330 | 0.199 | 0.262 | 0.286 | 0.261 | 0.418 | 0.354 | 0.306 |
| 7 | 0.232 | 0.233 | 0.308 | 0.234 | 0.334 | 0.232 | 0.361 | 0.324 | 0.377 | 0.220 |
| 8 | 0.191 | 0.209 | 0.226 | 0.140 | 0.324 | 0.259 | 0.216 | 0.220 | 0.339 | 0.265 |
| 9 | 0.321 | 0.169 | 0.290 | 0.230 | 0.229 | 0.297 | 0.235 | 0.218 | 0.203 | 0.348 |
| 10 | 0.312 | 0.173 | 0.167 | 0.322 | 0.177 | 0.259 | 0.137 | 0.267 | 0.258 | 0.268 |
| 11 | 0.277 | 0.291 | 0.287 | 0.199 | 0.192 | 0.275 | 0.311 | 0.318 | 0.275 | 0.307 |
| +gp | 0.277 | 0.291 | 0.287 | 0.199 | 0.192 | 0.275 | 0.311 | 0.318 | 0.275 | 0.307 |
| Fbar3-9 $^{0.270}$ | 0.239 | 0.308 | 0.231 | 0.295 | 0.261 | 0.313 | 0.293 | 0.319 | 0.303 |  |

Table 8.3.10 Sole VIIE Fishing Mortality at Age continued

| Age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | mean <br> F or-09 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 0.030 | 0.099 | 0.125 | 0.169 | 0.074 | 0.129 | 0.063 | 0.107 | 0.048 | 0.073 |
| 3 | 0.311 | 0.336 | 0.342 | 0.308 | 0.274 | 0.297 | 0.301 | 0.315 | 0.192 | 0.269 |
| 4 | 0.320 | 0.278 | 0.410 | 0.277 | 0.369 | 0.431 | 0.458 | 0.369 | 0.204 | 0.344 |
| 5 | 0.413 | 0.362 | 0.317 | 0.303 | 0.436 | 0.443 | 0.461 | 0.385 | 0.249 | 0.365 |
| 6 | 0.421 | 0.456 | 0.177 | 0.409 | 0.388 | 0.410 | 0.458 | 0.446 | 0.301 | 0.402 |
| 7 | 0.366 | 0.327 | 0.160 | 0.315 | 0.372 | 0.386 | 0.511 | 0.395 | 0.270 | 0.392 |
| 8 | 0.293 | 0.341 | 0.180 | 0.355 | 0.331 | 0.378 | 0.360 | 0.565 | 0.217 | 0.381 |
| 9 | 0.432 | 0.433 | 0.284 | 0.310 | 0.391 | 0.394 | 0.359 | 0.305 | 0.325 | 0.330 |
| 10 | 0.345 | 0.436 | 0.517 | 0.417 | 0.350 | 0.461 | 0.428 | 0.346 | 0.294 | 0.356 |
| 11 | 0.331 | 0.217 | 0.296 | 0.539 | 0.294 | 0.389 | 0.370 | 0.543 | 0.242 | 0.385 |
| +gp | 0.331 | 0.217 | 0.296 | 0.539 | 0.294 | 0.389 | 0.370 | 0.543 | 0.242 | 0.385 |
| Fbar3-9 $^{0.365}$ | 0.362 | 0.267 | 0.325 | 0.366 | 0.391 | 0.415 | 0.397 | 0.251 | 0.355 |  |

Table 8.3.11 Sole VIIE Summary Table

| Year | Recruits $\left[000{ }^{\prime}\right]$ | TSB $[\mathrm{t}]$ | SSB t$]$ | Landings $[\mathrm{t}]$ | Yield//SSB | FBar3-9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 1608 | 3338 | 2740 | 352.72 | 0.13 | 0.119 |
| 1970 | 3974 | 3599 | 3006 | 389.61 | 0.13 | 0.126 |
| 1971 | 2954 | 3285 | 2749 | 431.92 | 0.16 | 0.154 |
| 1972 | 2617 | 3557 | 2724 | 436.55 | 0.16 | 0.122 |
| 1973 | 3579 | 3894 | 3267 | 458.25 | 0.14 | 0.144 |
| 1974 | 3356 | 3999 | 3222 | 426.52 | 0.13 | 0.131 |
| 1975 | 3142 | 5136 | 4127 | 500.63 | 0.12 | 0.119 |
| 1976 | 7205 | 5584 | 4183 | 614.25 | 0.15 | 0.160 |
| 1977 | 5071 | 6018 | 4337 | 604.58 | 0.14 | 0.124 |
| 1978 | 4713 | 6571 | 4804 | 868.31 | 0.18 | 0.171 |
| 1979 | 5162 | 6746 | 5289 | 1170.17 | 0.22 | 0.229 |
| 1980 | 8942 | 6702 | 5206 | 1268.10 | 0.24 | 0.220 |
| 1981 | 5156 | 6332 | 4762 | 1217.81 | 0.26 | 0.260 |
| 1982 | 4164 | 6221 | 4728 | 1437.95 | 0.30 | 0.317 |
| 1983 | 6580 | 5888 | 4634 | 1503.84 | 0.32 | 0.367 |
| 1984 | 7835 | 5820 | 4548 | 1362.66 | 0.30 | 0.312 |
| 1985 | 4233 | 5968 | 3997 | 1400.09 | 0.35 | 0.343 |
| 1986 | 6445 | 5619 | 3899 | 1418.02 | 0.36 | 0.323 |
| 1987 | 4202 | 5415 | 3964 | 1279.28 | 0.32 | 0.293 |
| 1988 | 4113 | 5150 | 3875 | 1443.13 | 0.37 | 0.337 |
| 1989 | 3145 | 4509 | 3388 | 1389.36 | 0.41 | 0.413 |
| 1990 | 7902 | 5190 | 3242 | 1306.25 | 0.40 | 0.414 |
| 1991 | 4367 | 4476 | 2963 | 852.20 | 0.29 | 0.270 |
| 1992 | 3803 | 4179 | 2829 | 895.68 | 0.32 | 0.239 |
| 1993 | 2596 | 3606 | 2823 | 903.83 | 0.32 | 0.308 |
| 1994 | 3720 | 4225 | 3146 | 800.26 | 0.25 | 0.231 |
| 1995 | 4384 | 4464 | 3217 | 855.85 | 0.27 | 0.295 |
| 1996 | 3662 | 4640 | 3038 | 833.38 | 0.27 | 0.261 |
| 1997 | 4850 | 3778 | 2870 | 949.66 | 0.33 | 0.313 |
| 1998 | 3846 | 3936 | 2907 | 880.05 | 0.30 | 0.293 |
| 1999 | 7078 | 4910 | 2876 | 955.93 | 0.33 | 0.319 |
| 2000 | 5796 | 4897 | 2853 | 911.73 | 0.32 | 0.303 |
| 2001 | 4089 | 4479 | 2910 | 1068.62 | 0.37 | 0.365 |
| 2002 | 5922 | 4752 | 3058 | 1105.32 | 0.36 | 0.362 |
| 2003 | 2975 | 4396 | 3117 | 1078.12 | 0.35 | 0.267 |
| 2004 | 4225 | 4190 | 2895 | 1073.92 | 0.37 | 0.325 |
| 2005 | 4789 | 4253 | 2970 | 1036.77 | 0.35 | 0.366 |
| 2006 | 3811 | 3717 | 2519 | 1015.53 | 0.40 | 0.391 |
| 2007 | 3214 | 3695 | 2499 | 1014.65 | 0.41 | 0.415 |
| 2008 | 3037 | 3516 | 2277 | 908.12 | 0.40 | 0.397 |
| 2009 | $4332^{a}$ | 3552 | 2469 | 625.17 | 0.25 | 0.251 |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |

[^20]Table 8.3.12 Sole VIIE Short-term Forcast Input Table

2010

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 4332 | 0.10 | 0.00 | 0.00 | 0.00 | 0.091 | 0.000 | 0.124 |
| 2 | 3920 | 0.10 | 0.14 | 0.00 | 0.00 | 0.158 | 0.050 | 0.190 |
| 3 | 2371 | 0.10 | 0.45 | 0.00 | 0.00 | 0.221 | 0.185 | 0.251 |
| 4 | 1765 | 0.10 | 0.88 | 0.00 | 0.00 | 0.280 | 0.236 | 0.308 |
| 5 | 1428 | 0.10 | 0.98 | 0.00 | 0.00 | 0.335 | 0.251 | 0.361 |
| 6 | 1018 | 0.10 | 1.00 | 0.00 | 0.00 | 0.386 | 0.276 | 0.410 |
| 7 | 510 | 0.10 | 1.00 | 0.00 | 0.00 | 0.433 | 0.269 | 0.455 |
| 8 | 190 | 0.10 | 1.00 | 0.00 | 0.00 | 0.477 | 0.262 | 0.497 |
| 9 | 263 | 0.10 | 1.00 | 0.00 | 0.00 | 0.515 | 0.227 | 0.534 |
| 10 | 86 | 0.10 | 1.00 | 0.00 | 0.00 | 0.550 | 0.245 | 0.567 |
| 11 | 128 | 0.10 | 1.00 | 0.00 | 0.00 | 0.582 | 0.265 | 0.596 |
| 12 | 313 | 0.10 | 1.00 | 0.00 | 0.00 | 0.655 | 0.265 | 0.662 |

2011

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 4332 | 0.10 | 0.00 | 0.00 | 0.00 | 0.091 | 0.000 | 0.124 |
| 2 |  | 0.10 | 0.14 | 0.00 | 0.00 | 0.158 | 0.073 | 0.190 |
| 3 |  | 0.10 | 0.45 | 0.00 | 0.00 | 0.221 | 0.269 | 0.251 |
| 4 |  | 0.10 | 0.88 | 0.00 | 0.00 | 0.280 | 0.344 | 0.308 |
| 5 |  | 0.10 | 0.98 | 0.00 | 0.00 | 0.335 | 0.365 | 0.361 |
| 6 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.386 | 0.402 | 0.410 |
| 7 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.433 | 0.392 | 0.455 |
| 8 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.477 | 0.381 | 0.497 |
| 9 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.515 | 0.330 | 0.534 |
| 10 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.550 | 0.356 | 0.567 |
| 11 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.582 | 0.385 | 0.596 |
| 12 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.655 | 0.385 | 0.662 |

2012

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 4332 | 0.10 | 0.00 | 0.00 | 0.00 | 0.091 | 0.000 | 0.124 |
| 2 |  | 0.10 | 0.14 | 0.00 | 0.00 | 0.158 | 0.073 | 0.190 |
| 3 |  | 0.10 | 0.45 | 0.00 | 0.00 | 0.221 | 0.269 | 0.251 |
| 4 |  | 0.10 | 0.88 | 0.00 | 0.00 | 0.280 | 0.344 | 0.308 |
| 5 |  | 0.10 | 0.98 | 0.00 | 0.00 | 0.335 | 0.365 | 0.361 |
| 6 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.386 | 0.402 | 0.410 |
| 7 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.433 | 0.392 | 0.455 |
| 8 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.477 | 0.381 | 0.497 |
| 9 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.515 | 0.330 | 0.534 |
| 10 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.550 | 0.356 | 0.567 |
| 11 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.582 | 0.385 | 0.596 |
| 12 |  | 0.10 | 1.00 | 0.00 | 0.00 | 0.655 | 0.385 | 0.662 |

Table 8.3.13 Sole VIIE Single Option Output
$\underline{\text { Year }=2010 \mathrm{~F} \text {-multiplier }=0.688 \mathrm{Fbar}=0.244}$

| Age | F Catch No |  | Yield | Stock No | Biomass | SS No | SSB |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.000 | 0 | 0 | 4332 | 393 | 0 | 0 |
| 2 | 0.050 | 182 | 35 | 3920 | 618 | 549 | 87 |
| 3 | 0.185 | 382 | 96 | 2371 | 524 | 1067 | 236 |
| 4 | 0.236 | 354 | 109 | 1765 | 494 | 1554 | 435 |
| 5 | 0.251 | 302 | 109 | 1428 | 479 | 1400 | 469 |
| 6 | 0.276 | 234 | 96 | 1018 | 393 | 1018 | 393 |
| 7 | 0.269 | 115 | 52 | 510 | 221 | 510 | 221 |
| 8 | 0.262 | 42 | 21 | 190 | 91 | 190 | 91 |
| 9 | 0.227 | 51 | 27 | 263 | 135 | 263 | 135 |
| 10 | 0.245 | 18 | 10 | 86 | 47 | 86 | 47 |
| 11 | 0.265 | 28 | 17 | 128 | 75 | 128 | 75 |
| 12 | 0.265 | 69 | 46 | 313 | 205 | 313 | 205 |
| Total |  | 1778 | 618 | 16325 | 3675 | 7077 | 2394 |

$\underline{\text { Year }=2011 \text { F-multiplier }=1.000 \text { Fbar }=0.355}$

| Age | F Catch No |  | Yield | Stock No | Biomass | SS No | SSB |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.000 | 0 | 0 | 4332 | 393 | 0 | 0 |
| 2 | 0.073 | 262 | 50 | 3920 | 618 | 549 | 87 |
| 3 | 0.269 | 760 | 191 | 3374 | 746 | 1518 | 336 |
| 4 | 0.344 | 495 | 152 | 1783 | 499 | 1569 | 439 |
| 5 | 0.365 | 368 | 133 | 1261 | 423 | 1236 | 414 |
| 6 | 0.402 | 318 | 130 | 1006 | 389 | 1006 | 389 |
| 7 | 0.392 | 216 | 98 | 699 | 303 | 699 | 303 |
| 8 | 0.381 | 107 | 53 | 352 | 168 | 352 | 168 |
| 9 | 0.330 | 36 | 19 | 133 | 68 | 133 | 68 |
| 10 | 0.356 | 54 | 31 | 190 | 104 | 190 | 104 |
| 11 | 0.385 | 19 | 11 | 61 | 35 | 61 | 35 |
| 12 | 0.385 | 93 | 62 | 306 | 201 | 306 | 201 |
| Total |  | 2726 | 930 | 17416 | 3947 | 7618 | 2544 |

$\underline{\text { Year }=2012 \text { F-multiplier }=1.000 \text { Fbar }=0.355}$

| Age | F Catch No |  | Yield | Stock No | Biomass | SS No | SSB |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.000 | 0 | 0 | 4332 | 393 | 0 | 0 |
| 2 | 0.073 | 262 | 50 | 3920 | 618 | 549 | 87 |
| 3 | 0.269 | 743 | 186 | 3298 | 729 | 1484 | 328 |
| 4 | 0.344 | 647 | 199 | 2332 | 653 | 2052 | 575 |
| 5 | 0.365 | 334 | 120 | 1144 | 384 | 1121 | 376 |
| 6 | 0.402 | 250 | 103 | 792 | 306 | 792 | 306 |
| 7 | 0.392 | 188 | 86 | 609 | 264 | 609 | 264 |
| 8 | 0.381 | 129 | 64 | 427 | 204 | 427 | 204 |
| 9 | 0.330 | 58 | 31 | 218 | 112 | 218 | 112 |
| 10 | 0.356 | 25 | 14 | 86 | 47 | 86 | 47 |
| 11 | 0.385 | 37 | 22 | 120 | 70 | 120 | 70 |
| 12 | 0.385 | 69 | 46 | 226 | 148 | 226 | 148 |
| Total |  | 2742 | 921 | 17505 | 3928 | 7685 | 2516 |

input units are in 000 's and kg , output in t

Table 8.3.14 Sole VIIE Contributions and Source of Cohort for Short-term Forecast

| YC | Source | Yield2010 | Yield2011 | SSB2010 | SSB2011 | SSB2012 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2007 | XSA | 10.3 | 16.5 | 9.3 | 17.5 | 16.5 |
| 2008 | GM 69-07 | 3.8 | 21 | 3.4 | 14.7 | 23.3 |
| 2009 | GM 69-07 |  | 8 |  | 3.5 | 13.7 |
| 2010 | GM 69-07 |  |  |  |  | 3.4 |
| 2011 | GM 69-07 |  |  |  |  |  |

Cohort contributions to Yield2011


Cohort contributions to SSB2012


Table 8.3.15 Sole VIIE Management Options Output

| SSB | TSB | F-mult | F | basis | Yield <br> 2011 | SSB <br> 2012 | TSB <br> 2012 | $\%$ SSB- <br> Change | \%TAC- <br> Change |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2544 | 2011 |  |  |  | Fsq | 0 | 3406 | 4875 | 34 |
| 2544 | 3947 | 0.0 | 0.000 | 0.1 | 0.035 | Fsq | 107 | 3303 | 4765 |
| 2544 | 3947 | 0.2 | 0.071 | Fsq | 211 | 3203 | 4659 | 26 | -83 |
| 2544 | 3947 | 0.3 | 0.106 | Fsq | 312 | 3106 | 4557 | 22 | -66 |
| 2544 | 3947 | 0.4 | 0.142 | Fsq | 409 | 3013 | 4458 | 18 | -50 |
| 2544 | 3947 | 0.5 | 0.177 | Fsq | 503 | 2923 | 4362 | 15 | -19 |
| 2544 | 3947 | 0.529 | 0.188 | Fsq | 530 | 2898 | 4335 | 14 | -14 |
| 2544 | 3947 | 0.6 | 0.213 | Fsq | 594 | 2836 | 4269 | 11 | -4 |
| 2544 | 3947 | 0.627 | 0.222 | Fsq | 618 | 2813 | 4245 | 11 | 0 |
| 2544 | 3947 | 0.6761 | 0.240 | Fsq | 662 | 2772 | 4201 | 9 | 7 |
| 2544 | 3947 | 0.7 | 0.248 | Fsq | 682 | 2752 | 4180 | 8 | 10 |
| 2544 | 3947 | 0.7321 | 0.260 | Fsq | 710 | 2726 | 4151 | 7 | 15 |
| 2544 | 3947 | 0.8 | 0.284 | Fsq | 768 | 2671 | 4093 | 5 | 24 |
| 2544 | 3947 | 0.9 | 0.319 | Fsq | 850 | 2592 | 4009 | 2 | 38 |
| 2544 | 3947 | 1.0 | 0.355 | Fsq | 930 | 2516 | 3928 | -1 | 50 |
| 2544 | 3947 | 1.1 | 0.390 | Fsq | 1007 | 2443 | 3849 | -4 | 63 |
| 2544 | 3947 | 1.2 | 0.425 | Fsq | 1082 | 2372 | 3773 | -7 | 75 |
| 2544 | 3947 | 1.3 | 0.461 | Fsq | 1155 | 2304 | 3700 | -9 | 87 |
| 2544 | 3947 | 1.4 | 0.496 | Fsq | 1225 | 2238 | 3628 | -12 | 98 |
| 2544 | 3947 | 1.5 | 0.532 | Fsq | 1293 | 2174 | 3560 | -15 | 109 |
| 2544 | 3947 | 1.6 | 0.567 | Fsq | 1358 | 2112 | 3493 | -17 | 120 |
| 2544 | 3947 | 1.7 | 0.603 | Fsq | 1422 | 2052 | 3428 | -19 | 130 |
| 2544 | 3947 | 1.8 | 0.638 | Fsq | 1484 | 1994 | 3366 | -22 | 140 |
| 2544 | 3947 | 1.9 | 0.674 | Fsq | 1543 | 1938 | 3305 | -24 | 150 |
| 2544 | 3947 | 2.0 | 0.709 | Fsq | 1601 | 1884 | 3247 | -26 | 159 |
| 2544 | 3947 | 0.762 | 0.270 | Fmsy | 736 | 2701 | 4125 | 6 | 19 |
| 2544 | 3947 | 0.85 | 0.301 | Fmp F | 809 | 2631 | 4051 | 3 | 31 |
| 2544 | 3947 | 0.7321 | 0.260 | Fmp TAC | 710 | 2726 | 4151 | 7 | 15 |

Table 8.3.16: Sole VIIe senfile used as input data in MSY analysis

|  | 9 | 2009 | 3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 |  |  |  |
| 'N1' | 2815 | 0.38 | continued |  |  |
| 'N2' | 4753 | 0.6 | 'M1' | 0.1 | 0.1 |
| 'N3' | 2371 | 0.49 | 'M2' | 0.1 | 0.1 |
| 'N4' | 1765 | 0.21 | 'M3' | 0.1 | 0.1 |
| 'N5' | 1428 | 0.19 | 'M4' | 0.1 | 0.1 |
| 'N6' | 1018 | 0.16 | 'M5' | 0.1 | 0.1 |
| 'N7' | 510 | 0.15 | 'M6' | 0.1 | 0.1 |
| 'N8' | 190 | 0.15 | 'M7' | 0.1 | 0.1 |
| 'N9' | 263 | 0.14 | 'M8' | 0.1 | 0.1 |
| 'N10' | 86 | 0.15 | 'M9' | 0.1 | 0.1 |
| 'N11' | 128 | 0.13 | 'M10' | 0.1 | 0.1 |
| 'N12' | 313 | 0.13 | 'M11' | 0.1 | 0.1 |
| 'sH1' | 0 | 0 | 'M12' | 0.1 | 0.1 |
| 'sH2' | 0.073 | 0.587 | 'MT1' | 0 | 0 |
| 'sH3' | 0.269 | 0.107 | 'MT2' | 0.14 | 0.1 |
| 'sH4' | 0.344 | 0.125 | 'MT3' | 0.45 | 0.1 |
| 'sH5' | 0.365 | 0.096 | 'MT4' | 0.88 | 0.1 |
| 'sH6' | 0.402 | 0.074 | 'MT5' | 0.98 | 0 |
| 'sH7' | 0.392 | 0.092 | 'MT6' | 1 | 0 |
| 'sH8' | 0.381 | 0.214 | 'MT7' | 1 | 0 |
| 'sH9' | 0.33 | 0.181 | 'MT8' | 1 | 0 |
| 'sH10' | 0.356 | 0.148 | 'MT9' | 1 | 0 |
| 'sH11' | 0.385 | 0.327 | 'MT10' | 1 | 0 |
| 'sH12' | 0.385 | 0.327 | 'MT11' | 1 | 0 |
| 'WH1' | 0.121 | 0.126 | 'MT12' | 1 | 0 |
| 'WH2' | 0.185 | 0.065 | 'R11' | 4332 | 0.35 |
| 'WH3' | 0.242 | 0.048 | 'R12' | 4332 | 0.35 |
| 'WH4' | 0.297 | 0.054 | 'HF10' | 1 | 0.1 |
| 'WH5' | 0.348 | 0.06 | 'HF11' | 1 | 0.1 |
| 'WH6' | 0.397 | 0.064 | 'HF12' | 1 | 0.1 |
| 'WH7' | 0.443 | 0.066 | 'K10' | 1 | 0.1 |
| 'WH8' | 0.486 | 0.068 | 'K11' | 1 | 0.1 |
| 'WH9' | 0.526 | 0.07 | 'K12' | 1 | 0.1 |
| 'WH10' | 0.562 | 0.072 | Sole |  |  |
| 'WH11' | 0.597 | 0.076 | VIle |  |  |
| 'WH12' | 0.687 | 0.079 | 1 |  |  |
| 'WS1' | 0.08 | 0.419 | 1 | 12 | 1 |
| 'WS2' | 0.155 | 0.095 | 1 |  |  |
| 'WS3' | 0.214 | 0.052 | H.cons. |  |  |
| 'WS4' | 0.27 | 0.051 | 3 | 9 |  |
| 'WS5' | 0.323 | 0.057 | 1969 | 2009 |  |
| 'WS6' | 0.373 | 0.062 | Stock num | in 2010 | VPA |
| 'WS7' | 0.42 | 0.065 | -1 |  |  |
| 'WS8' | 0.465 | 0.067 |  |  |  |
| 'WS9' | 0.506 | 0.069 |  |  |  |
| 'WS10' | 0.544 | 0.07 |  |  |  |
| 'WS11' | 0.58 | 0.074 |  |  |  |
| 'WS12' | 0.675 | 0.074 |  |  |  |

Table 8.3.17: Sole in Division VIle : Estimates of biomass and fishing mortality reference levels derived from the fit of three stock and recruit relationships and the yield per recruit Fmsy proxies.
Stock name
Sole VIIe
Sen filename
wgcse_sol7e.sen
pf, pm
$0 \quad 0$
Number of iterations
1000
Simulate variation in Biological parameters
TRUE
SR relationship constrained
TRUE
Ricker
905/1000 Iterations resulted in feasible parameter estimates

|  | Fcrast | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 0.335 | 0.144 | 9994 | 1475.9 | 1.038 | 0.200 | 1.678 | 0.000073 | 13.926 |
| Mean | 0.341 | 0.145 | 15146 | 2001.7 | 1.050 | 0.236 | 1.784 | 0.000086 |  |
| 5\%ile | 0.210 | 0.093 | 4442 | 702.19 | 0.967 | 0.049 | 1.373 | 0.000018 |  |
| 25\%ile | 0.275 | 0.119 | 6337 | 1034.3 | 1.008 | 0.138 | 1.557 | 0.000050 |  |
| 50\%ile | 0.328 | 0.140 | 9088 | 1343.7 | 1.046 | 0.228 | 1.730 | 0.000083 |  |
| 75\%ile | 0.397 | 0.165 | 14964 | 2022.3 | 1.088 | 0.320 | 1.962 | 0.000117 |  |
| 95\%ile | 0.521 | 0.213 | 41855 | 5119.2 | 1.152 | 0.467 | 2.352 | 0.000170 |  |
| CV | 0.284 | 0.248 | 1.538 | 1.313 | 0.055 | 0.543 | 0.175 | 0.543 |  |

## Beverton-Holt

905/1000 Iterations resulted in feasible parameter estimates

|  | Fcrast | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Unscaled Beta AIC

## Smooth hockeystick

905/1000 Iterations resulted in feasible parameter estimates

|  | Fcrash | Fmsy | Bmsy | MSY | ADMB Alpha | ADMB Beta | Unscaled Alpha | Unscaled Beta | AIC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Determinis | 0.263 | 0.263 | 4220 | 1143 | 0.505 | 1.227 | 0.668 | 4217.600 | 15.668 |
| Mean | 0.253 | 0.216 | 8393 | 1173 | 0.507 | 1.279 | 0.672 | 4395.680 |  |
| 5\%ile | 0.164 | 0.039 | 3613 | 783 | 0.464 | 1.034 | 0.614 | 3554.802 |  |
| 25\%ile | 0.210 | 0.182 | 4158 | 982 | 0.488 | 1.168 | 0.646 | 4015.100 |  |
| 50\%ile | 0.250 | 0.225 | 4628 | 1141 | 0.505 | 1.277 | 0.669 | 4391.610 |  |
| 75\%ile | 0.290 | 0.264 | 5188 | 1338 | 0.523 | 1.414 | 0.693 | 4860.090 |  |
| 95\%ile | 0.354 | 0.326 | 35746 | 1643 | 0.551 | 1.514 | 0.729 | 5206.500 |  |
| CV | 0.227 | 0.368 | 1.462 | 0.230 | 0.059 | 0.122 | 0.059 | 0.122 |  |

Yield per Recruit

|  | F35 | F40 | F01 | Fmax | Bmsypr | MSYpr | Fpa | Flim |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Determinis | 0.142 | 0.118 | 0.142 | 0.376 | 0.749 | 0.203 | 0.2 | 0.28 |
| Mean | 0.131 | 0.109 | 0.129 | 0.422 | 1.424 | 0.200 |  |  |
| 5\%ile | 0.002 | 0.002 | 0.002 | 0.039 | 0.692 | 0.141 |  |  |
| 25\%ile | 0.066 | 0.054 | 0.073 | 0.261 | 0.732 | 0.171 |  |  |
| 50\%ile | 0.141 | 0.117 | 0.144 | 0.375 | 0.765 | 0.195 |  |  |
| 75\%ile | 0.191 | 0.159 | 0.185 | 0.501 | 0.817 | 0.225 |  |  |
| 95\%ile | 0.253 | 0.209 | 0.235 | 0.932 | 6.052 | 0.274 |  |  |
| CV | 0.624 | 0.626 | 0.594 | 0.765 | 1.436 | 0.205 |  |  |

Figure 8.3.1 Sole VIIE International Landings Age Compositions


Figure 8.3.2 Sole VIIE Catch and Stock Weights at Age

## Catch Weights for Sole VIIE (age 1 to 12+)



## Stock Weights for Sole VIIE (age 1 to 12+)



Figure 8.3.3a Sole VIIE Discards by Quarter, Fleet


Figure 8.3.3b Sole VIIE Discards by Quarter, Fleet continued

## FRTrawl




Figure 8.3.3c Sole VIIE Discards by Quarter, Fleet continued

## FRNets




Figure 8.3.4 Sole VIIE LPUE and effort


Means standardiused effort for UK Beamtrawlers


Figure 8.3.5 Sole VIIE Log CPUE by Yearclass
note the cohorts differ on the x-axes due to the differences in the length and age range of the tuning series




Figure 8.3.6 Sole VIIE Log CPUE by Year
note the cohorts differ on the x-axes due to the differences in the length and age range of the tuning series


Figure 8.3.7 Sole VIIE Single Fleet log catchability Residuals

UK-CBT





Figure 8.3.8 Sole VIIE Single Fleet Summary



Recruitment


Figure 8.3.9 Sole VIIE Final XSA Fleet log catchability Residuals












Figure 8.3.10 Sole VIIE Final XSA and previous XSAs
Fishing Mortality




Figure 8.3.11 Sole VIIE Final and previous Assessment weights




Figure 8.3.12 Sole VIIE XSA Retrospective Plots
Fishing Mortality




Figure 8.3.13: Sole in Divisions VIIe : Fitted yield per recruit F reference points, yield per recruit and SSB per recruit against fishing mortality with confidence intervals estimated by parametric re-sampling of the selection, weight at age, natural mortality and maturity estimates and their c.v. Left hand panels : blue line indicates the deterministic estimate, red lines the median and percentiles. Right hand panels : the first 100 re-samples.


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## Annex 2: Technical minutes from the Celtic Seas Review Group

- RGCS
- By correspondence and through WebEx conference 26-28 May 2010
- Participants: Mike Armstrong (Chair), Marie Storr-Paulson, Jens Floeter, Yvonne Walther. WG Chairs: Pieter-Jan Schön and Colm Lordan, Ireland. Secretariat: Cristina Morgado, Barbara Schoute, Mette Bertelsen and Helle Gjeding Jørgensen.
- Working Group: WGCSE


## General

## Stocks to be reviewed

The Review Group considered the following stocks:

| Fish stocks | Perform assessment | Advice |
| :---: | :---: | :---: |
| Anglerfish (Lophius piscatorius and L. budegassa) in Division IIa, IIIa, Subarea IV and VI | Y | Update |
| Cod in Division VIIe-k (Celtic Sea) | Y | Update |
| Cod in Division VIIa (Irish Sea) | Y | Update |
| Cod in Division VIb (Rockall) | N | Catch statistics only |
| Cod in Division VIa (West of Scotland) | Y | Update |
| Haddock in Divisions VIIb-k | Y | Update |
| Haddock in Division VIIa (Irish Sea) | Y | Update |
| Haddock in Division VIb (Rockall) | Y | Update |
| Haddock in Division VIa (West of Scotland) | Y | Update |
| Megrim (Lepidorhombus spp) in Subarea VI (West of Scotland and Rockall) and Subarea IV (North Sea) | Y | Update |
| Nephrops in Division VIa (North Minch) | Y | Update |
| Nephrops in Division VIa (South Minch) | Y | Update |
| Nephrops in Division VIa (Firth of Clyde) | Y | Update |
| Nephrops in Division VIIa (Irish Sea East) | Y | Update |
| Nephrops in Division VIIa (Irish Sea West) | Y | Update |
| Nephrops in Division VIIb, c,j,k (Porcupine Bank) | Y | Update |
| Nephrops in Division VIIb (Aran Grounds, FU 17) | Y | Update |
| Nephrops in Division VIIa,g,j (South East and West of IRL, FU 19) | Y | Update |
| Nephrops in Divisions VIIfgh (Celtic Sea, FU 20-22 | Y | Update |
| Plaice in Division VIIb,c (West of Ireland) | Y | Update |
| Plaice in Divisions VIIh,k (Southwest of Ireland) | Y | Same advice as last year |
| Plaice in Divisions VIIf,g (Celtic Sea) | Y | Update |
| Plaice in Division VIIe (Western Channel) | Y | Update |
| Plaice in Division VIIa (Irish Sea) | Y | Update |
| Sole in Division VIIb, c (West of Ireland) | N | Catch statistics only |
| Sole in Divisions VIIh-k (Southwest of Ireland) | Y | Same advice as last year |


| Fish stocks | Perform <br> assessment | Advice |
| :--- | :---: | :--- |
| Sole in Divisions VIIf,g (Celtic Sea) | Y | Update |
| Sole in Division VIIe (Western Channel) | Y | Update |
| Sole in Division VIIa (Irish Sea) | Y | Update |
| Whiting in Divisions VIIe-k | Y | Update |
| Whiting in Division VIIa (Irish Sea) | Y | Update |
| Whiting in Division VIb (Rockall) | N | Catch statistics only |
| Whiting in Division VIa (West of Scotland) | Y | Update |

The following special requests were addressed:

## Additional Requests to ICES concerning West Scotland haddock

On behalf of the European Commission, ICES is requested and required to incorporate the following elements in its advice to be provided in June 2010.
a) ICES should report on the catch of haddock in Divisions Vb and VIa in 2010 that would be consistent with the application of the harvest rule in Annex, paragraph 1.

Such reports could adequately be presented as additional lines in the "catch option table".

## Special request on VIIe sole

Request from UK authorities to consider potential alternative methods for estimating the current exploitation rate on VIIe sole and to identify an appropriate method for the calculation of a TAC for 2011, in accordance with the management plan.

## Review process

The Review Group conducted its work by correspondence and through Webex conference facilities organised by ICES. The reviews have been carried out according the Guidelines provided by ICES, particularly focusing on the need to Quality Assure the assessment results supporting the provision of fishery management advice by ICES in the annual ACOM advice sheets. All stocks were reviewed by at least two reviewers. This involved:

- Checking that update assessments have been correctly implemented using the methods described in the Stock Annexes;
- Checking that the assessments have been implemented correctly, which could involve re-running the assessments to ensure the results in the WG Report can be replicated exactly;
- Ensuring the assessment results and forecast results are carried over correctly to the advice sheets and advising ICES of any errors detected;
- Evaluating the ability of the stock assessments for providing credible management advice, and suggesting alternative advice where assessments do not appear appropriate;
- Providing recommendations to the Working Group to help with future development of the assessments through benchmarking.

The RG did not have access to all of the WG report sections and advice sheets prior to the scheduled start date of the review process on 26 May. This meant that a very heavy workload was experienced during the week commencing 31 May. Unfortunately one of the three reviewers became unavailable during that week, increasing
the workload on the remaining reviewers. The RG advises that insufficient time was provided to review the large number of stocks, which included four herring stocks from HAWG and anglerfish and megrim stocks from WGHMM. ICES must review the logistics of the review process for future years.

## General comments regarding the WGCSE stock assessments

Several issues were common to many of the assessments carried out by the WGCSE:

## 1 ) Discard estimates

Many of the stocks are heavily discarded at the younger age classes and also in some cases throughout the age classes due to high-grading. Apart from west-of-Scotland gadoids, there is an absence of appropriately fleet-raised estimates of discards for all the significant fleets included in the assessments (Scotland is currently reviewing its raising procedures). In some cases (e.g. plaice in VIIa and VIIfg; whiting in VIIa) where very high rates of discarding occur across many age classes but fleet-raised estimates are not provided to the WG by all countries, there is no valid basis for a catch-at-age assessment and any estimates of fishing mortality from a landings-only assessment will be severely biased. EU member States are required through the Data Collection Framework to collect data on discards for fleets where discarding exceeds a specified percentage, and are expected to meet precision targets. This has been a requirement since the inception of the DCR/DCF. However, discards estimates are often not being transmitted to the WG in a usable form for inclusion in an assessment, or in some cases are based on extremely low sample sizes. In some cases it is not even possible to quantify the general level of discarding.

2 ) Biological sampling on surveys
Considerable archives of biological data collected on surveys exist for many stocks. However, very little of this is used by the WG to provide time-series of biological parameters such as maturity, length/weight-at-age, etc. For example, the practice of using mean weights-at-age in annual commercial catches (or even worse, in landings) as values for weight-at-age in the stock, is prevalent despite the existence of good data from surveys at different times of year, which can be modelled to obtain year and age or year-class effects (see VIIa haddock for example). The benchmark assessment process provides a focus for such analyses, but the WGCSE stocks have been poorly serviced by benchmarks so far.

3 ) Commercial lpue tuning data
There are a number of stocks (typically flatfish) where commercial fleet tuning data are still used in the assessments. In some cases these comprise a large fraction of the total fishery catch and this leads to correlated errors between the tuning fleets and the catch-at-age matrix. The Stock Annexes do not always provide adequate evidence to support the contention of constant catchability over time, for example contemporary evaluation of "power factors", and evaluation of how catchability is affected by changes in fleet behaviour caused by management regimes or other drivers such as fuel costs. It is not uncommon for surveys and commercial tuning fleets to give conflicting signals. The benchmark process should involve full evaluation of commercial tuning fleets if these are to be used. VMS data linked to EU logbook data, and observer data, provide an opportunity to examine spatio-temporal patterns in cpue, at least for the larger vessels.

4 ) $\mathrm{F}_{\mathrm{MSY}}$ estimates

The RG appreciates the efforts of the WG to explore candidate Fmsy values for the different stocks. However there is a clear problem with characterising the true nature of the stock-recruit relationship, which is critical for robust estimation of Fmsy. Problems exist due to the characteristics of the variability in recruitment, the lack of any apparent decline in recruitment at reduced SSB, and time-series effects suggesting a Rickertype domed $S / R$ even when the time-series covers a range of stock sizes that are already well depleted from the "virgin" stock. Bootstrapping from often noisy assessment data also leads to difficulties in finding solutions, making it difficult even to make robust estimates of yield-per-recruit parameters in some cases. A basic problem is that individual stocks cannot yield sufficient statistical power to define the parameters due to natural variability and sampling error. As a result, the WG has been unable to specify $\mathrm{F}_{\mathrm{mSY}}$ targets for many stocks, which is a problem given the commitments to an MSY approach to management. It seems more likely that a metaanalysis approach would lead to more robust outcomes; there are likely to be similarities in appropriate $\mathrm{Fmsy}_{\text {m refere }}$ refere points between stocks with similar biology and dynamics, and the variability between the "true" values for these stocks is probably much less than the estimation errors and biases for individual assessments.

5 ) Benchmarking
So far, few of the stocks covered by WGCSE have been the subject of benchmark assessments. Without the target of a data compilation and benchmark assessment, national scientists may not have the leverage to find the resources to carry out the necessary intersessional preparatory work. This can lead to an argument that the stocks are not ready for benchmarking because the data have not been adequately compiled, resulting in perpetuation of inadequate assessments and advice. All efforts should be made to benchmark stocks that have important linkages, e.g. gadoid stocks in the North Sea and west of Scotland, or western waters cod stocks.

6 ) Quality of official landings data
The non-provision of official landings data by France in 2009 has affected the quality of landings dataseries for many stocks in this and other WGs. In many cases French data have been obtained from a different source, e.g. sales slips, but it is not clear how accurate these estimates are.

7 ) Stock Annexes
A number of Stock Annexes are comprehensively updated each year but others contain old information or data evaluations (e.g. catch-curves) which haven't been updated and are of limited current value. The Annexes should be brought fully up to date and any out-of-date analyses revised if they are still thought to be useful, or removed if no longer of use and aren't necessary for interpreting historical data or assessments.

8 ) Nephrops assessments
The assessments and advice for most of the Nephrops stocks in Subareas VI and VII have fallen into a common approach based on UWTV surveys, yield-per-recruit FmsY estimates, and supporting trends indicators. The RG appreciates the efforts by WGCSE to standardise the methods and streamline the reports. However there remain some inconsistencies in the way in which key variables are presented. The sections for the Aran and Irish Sea west stocks include useful tables detailing all the inputs for calculating historical harvest ratios and inputs to forecasts, but other stocks lack this detail making it difficult for the RG to evaluate the assessments. This was
not helped by errors in some key tables for west-of-Scotland stocks. Efforts to standardise the tables further would be appreciated.

## Anglerfish (Lophius piscatorius and L. budegassa) in Division IIa, IIIa, Subarea IV and VI (Report Section 5.2)

1) Assessment type: Update

2 ) Assessment: trends
3 ) Forecast: none
4 ) Assessment model: No assessment model presented
5 ) Consistency: Survey-series 2005-2008 revised
6 ) Stock status: Unknown
7 ) Management Plan: There is no management plan for the stock

## General comments

The WG has addressed the ToR in providing an update of survey data and other indicators.

The WG Report and the Stock Annex do not contain any information on ecosystem aspects or environmental drivers.

There is no EU management plan for this stock.
A detailed description of the anglerfish fisheries is given in the Stock Annex. However the WG has not considered mixed fishery issues in relation to management of anglerfish in its advice.

A new revised survey estimate from the joint science/industry anglerfish survey was introduced. The 2010 survey data has not been finalised yet. The 2010 survey result could be used for evaluating TAC options for 2011 using the Commission rules in its Consultation on Fishing Opportunities for 2010, and a final recommendation should be given as along with other ICES' survey updates later on in the year. The survey time-series goes back to 2005 .

There are age reading problems for the species a workshop is planned in 2011. There is concern of area and species misreporting.

A section with management consideration for the IIIa, IV and VI stock is missing and it is unclear what management advice the WG is providing.

## Technical comments

1) The collaborative survey-series has been revised this year. The WG should provide a comparison with the results from last year's WG and explain any differences.

2 ) In the MSY section the WG provides an approximate estimate of $Z=0.6$ from the survey catch-at-age data but does not provide the diagnostics of the method including how selectivity is accounted for.

3 ) The increasing trend in the anglerfish survey catch rates is also reflected in the lpue in the Irish fishery in VIb (Rockall). However the survey and the Irish lpue have opposite trends in VIa. The WG has not provided updates of Scottish lpue (using days fished as effort) due to historical inaccuracies in landings data. However it is stated that the reported landings data from 2006 onwards are more accurate due to Buyers and Sellers legislation. Spatial lpue trends from 2006 on would be useful additional evidence given the downward trend in the survey index in 2009.

## Conclusions

The RG agrenes with the conclusions of the WG and the material in the draft advice sheets that no reliable assessment can be provided at this stage.

The stock suffers from data deficiencies regarding age reading, area misreporting and limited knowledge about population dynamics. The new joint science/industry anglerfish survey is an important initiative for providing biomass estimates and stock trends. Catchability with respect to the efficiency of the trawl in retaining anglerfish in the path of the net has been studied in detail although there are still catchability issues related to distribution (for example underrepresentation of young anglerfish). Survey data indicate a sharp decrease in numbers from 2007-2009 and a sharp drop in biomass from 2008-2009. However the time-series is very short and the propensity for year effects in survey results is not apparent yet.

## Cod in Division VIle-k (Celtic Sea) (Report Section 7.2)

1 ) Assessment type: Update
2 ) Assessment: survey and fishery trends
3 ) Forecast: none provided
4 ) Assessment model: none
5 ) Consistency: failed benchmarked in 2009, no new assessment proposed
6 ) Stock status: Unknown. Yield-per-recruit analysis suggests $\mathrm{F}_{0.1}$ (0.27) and $F_{\max }$ (0.38) values that are consistent with previous analyses, and well below historical F reported in previous XSA.
7 ) Management Plan: No agreed management plan has been developed yet. However, a long-term management plan is under discussion for this stock and an effort based management system in the Celtic Sea (VIIfg) is being discussed by member states and the EC.

## General comments

The specific ToR for VIIe-k cod was to perform an update assessment (as opposed to SALY). The WG continues to follow the WKROUND advice not to perform an analytical assessment due to catch uncertainties. This unfortunately precludes any presentation of long-term trends in SSB, F and recruitment other than the separable VPA recruitment-series presented, and it is not possible to see if the addition of new data has affected the WKROUND conclusions. This leaves a critical cod stock with very little quantitative advice on stock status.

The WG Report and Stock Annex do not include any ecosystem information relevant to this stock and its fisheries, or any information on climate changes that could affect the stock. Given the location of the stock at the southern limits of the species range, this is a major omission that the WG should address. The Stock Annex comment that "no environmental drivers are known for this stock" is not correct given the history of studies on cod and climate.

There is no agreed Management Plan for this stock.
The WG has not used mixed fishery data in the stock section or annex other than a statement of which fisheries cod is caught in. Cod in the Celtic Sea are often taken as a minor bycatch in a range of trawl and netting fisheries targeting a diverse range of species. Management measures to conserve cod could impact a wide range of fisheries that do not target cod and the impact of this, needs to be evaluated.

As with other stocks, official landings data were not received from France for 2009 and catches had to be derived from landings reported by fishing organizations. Only $72 \%$ of the French quota was taken, resulting in an apparent sharp reduction in dis-carding/high-grading in 2009. Discarding (mainly high-grading with some undersized cod) and landings misreporting had been exacerbated since 2003, when quotas became increasingly restrictive. Discard data were presented in previous WGSSDS but not used in the assessments as they did not cover all the main fleets and quarters yet.

The two main problems in assessing the stock are: 1) perceived problems with accuracy of catch-at-age data in the 2000s due to high-grading (although attempts have been made to adjust for this), and (2) available surveys lack robust trends mainly due to their low catch rates, with all current survey-series taking place in autumn when cod are dispersed and often present in non-trawlable grounds. A new Irish industry
led survey during spring is commencing, which should provide important new data on distribution and abundance but will require several years to indicate stock trends.

## Technical comments

1 ) SURBA results are given in terms of $Z$, but only the raw indices of recruitment for the EVHOE survey are given for comparison with separable VPA. Given the noise in the survey data, the use of SURBA recruitment estimates (using several data points for some year classes) may be more robust and should be investigated.

2 ) The WG makes a strong statement that the range of the stock has contracted, based on the plots in Figs 7.2.6-7.2.8. It is not clear if this is the case (the yellow areas cover the same area each year) or if what is being seen is a general reduction in the overall catch level across the stock range.
3 ) The Report contains many tables of catch-at-age and length frequencies. It would be useful to highlight the ages and years where high-grading is expected to be sufficient to have introduced an unacceptable bias in the XSA. The main question for the WG to address is how the bias/imprecision in these catch-at-age values can be accounted-for in a suitable analytical model formulation.
4 ) The WG should provide a better explanation of the YPR inputs, including why the stock weights and spawning-stock weights are different over all the ages.

5 ) An "exploratory VPA" is mentioned as the source of YPR selectivity data, and the recruit trends from a separable VPA are given elsewhere. However no details of this VPA are given.
6 ) In the Stock Annex two commercial French tuning fleets are described: the French trawlers targeting Gadoids in Divisions VIIf, g, h (FR-GADOIDS) and the French Nephrops trawlers in VIIf,g,h (FR-NEPHROPS). The Q2-Q4 data for these were used in previous assessments, but the series have not been updated in the WG Report, presumably due to the lack of French data for 2009?

## Conclusions

The RG agrees that this stock requires a further benchmark assessment to review and take into account more recent information. However, the proposed date in late 2011 will mean no new advice on this stock until 2012, relating to TACs for 2013.

The RG supports the continued efforts to improve the input data, including the instigation of the new Q1 survey by the Marine Institute and the Federation of Irish Fishermen in 2010.

## Cod in Division VIla (Irish Sea) (Report Section 6.2)

1 ) Assessment type: Update
2 ) Assessment: analytical
3 ) Forecast: a B-adapt short-term predictions
4 ) Assessment model: B-Adapt-F.exe (13/5/06) with 5 survey indices
5 ) Consistency: very consistent with last year
6 ) Stock status: The spawning-stock biomass has declined ten-fold since the late 1980s and is suffering reduced reproductive capacity (SSB2009 =1192t) <Blim of 6000 t ). The fishing mortality estimates since 1988 have remained above the $\mathrm{F}_{\text {lim }}$ value of $\mathrm{F}=1.0$ and the stock has therefore been harvested unsustainably over this period.
7 ) Management Plan: In 2008 the EU adopted a long-term plan for cod stocks and the fisheries exploiting those stocks (Council Regulation (EC) 1342/2008) that repeals Regulation (EC) No 423/2004, and has the objective of ensuring the sustainable exploitation of the cod stocks on the basis of maximum sustainable yield while maintaining a target fishing mortality of 0.4 on specified age groups.

## General comments

The Report is well written and has a good structure. The WG addressed the ToRs by providing an update with associated management advice.

The assessment was carried out according to the Stock Annex description. The RG found no errors in the implementation of the assessment, and the results were carried over correctly to the advice sheets.

The WG Report and Stock Annex do not include any ecosystem information. Previous WGs (2006) documented a relationship between temperature in spring and cod recruitment anomalies and this is mentioned in the Annex. However the assessment and advice do not take climate effects into account. The WG Report includes information on stock structure and migrations based on tagging studies, indicating mixing of cod stocks between the Irish Sea, west of Scotland and the Celtic Sea.

The stock is subject to an EU multi-annual management plan. ICES (2009 Advice) evaluated the management plan, and considers the implementation to Irish Sea cod is not in accordance with the precautionary approach if a constraint on interannual TAC adjustments is applied, given the poor state of the stock.

The WG has not used mixed fishery data in providing management advice.
Discards data are not included in the assessment. Data presented up to 2009 show discarding on the observer trips to have been predominantly undersized cod at ages 0 and 1, rather than the high-grading observed in some other cod stocks. The RG noted that discards data for UK (E\&W) are presented by gear up to 2006 but for combined gears thereafter.

The major issue with this assessment is the apparently large "catch bias" estimated by B-Adapt over the 2000-2009 period. Although slowly reducing, the bias multiplier remains between 2 and 3 . This represents the multiplier that has to be applied to the landings-at-age to remove any catchability trends in the surveys. The WG still has no knowledge of the cause of such removals, or the extent to which the bias may reflect true catchability trends in the surveys. The TAC has not been fully utilized for a
number of years, and landings reporting is considered more accurate since introduction of Buyers and Sellers legislation in 2006.

The WG Irish Sea Overview Section was not available to the RG to evaluate effort trends. Last year's WG showed large declines in the whitefish trawl effort ( $100 \mathrm{~mm}+$ mesh) during the 2000s, and more stable Nephrops trawl effort. It is difficult to reconcile this with the lack of any recovery in age composition as shown by the continued very high Z . All available trawl data indicate a truncated age distribution in the Irish Sea. The declining SSB is supported by the Fishery Science Partnership surveys not included in the assessment.

## Technical comments

1 ) The three 0-gp indices in the assessment indicated above-average 2004 and 2005 year classes that were not evident in the March GFS or the B-Adapt. The indices for the 2009 year class are divergent, but the large 1-gp index in the March 2010 survey gives more confidence in the possibility of an above average year class. The decision to conduct medium-term forecasts with and without the 2009 estimate is sensible. However the results do not change the management advice.
2 ) In the Stock Annex the survey ages for ScoGFS-Q1 Survey is 1-5, but in the input data it is age 1-4 (however, the time-series stopped in 2006 and the difference is probably very small).
3 ) Table 6.2.8 diagnostics. For the survey NIGFS_oct ( $0-2 \mathrm{gr}$ ) the t -value is very high 3.79 indicating a density-dependent catchability for this age. Future benchmarks should consider how 0-gp trawl indices are handled in the assessment, as the preferred habitat for 0 -gp cod would be inshore rough ground poorly covered by the surveys.
4 ) MSY evaluation: The WG used only data up to 2005 for F due to the uncertainty in the F estimation for the most recent years in this stock. The retrospectives show some large adjustments to F (e.g. in 2005 and 2007), and the large 2006 F estimate is probably effectively converged. It would be more useful to examine the effect on the $\mathrm{F}_{\text {mSy }}$ estimates of different year ranges of assessment results. The catch bias is also highest in 2003 and has been declining.

## Advice sheets

1) The reference points table in the advice sheet has the wrong Fmsy range (0.24-0.49). The correct values are $0.25-0.54$. This error was transmitted to ICES.

## Conclusions

The RG accepts the updated assessment as a basis for providing advice on the state of the stock relative to biological reference points.

The RG recommends that the assessment of this stock should be benchmarked in the context of a benchmarking of all three western waters cod stocks, given the availability of new data not included in the assessments (FSP surveys; egg production survey estimates; discards data collected in the 2000s through the DCF) and improving knowledge of metapopulation structure and movements of cod in the overall ecoregion. Other issues for Irish Sea cod that could be addressed include:

- Modelling framework that is not constrained to estimate equal catch-biases for all ages and fleets and can avoid the need to truncate the oldest true age to 4 for the entire time-series;
- Modelling of survey data to generate stock weights and maturity ogives (weights-at-age in the catches are becoming very variable at the older ages due to small sample sizes).


## Cod in Division VIb (Rockall) (Report Section 4.2)

1 ) Assessment type: No advice; catch statistics only
2 ) Assessment: not presented
3 ) Forecast: not presented
4 ) Assessment model: none
5 ) Consistency:
6 ) Stock status: unknown
7 ) Management Plan: none

## General comments

Official landing are below 100 t decreasing from close to 2000 t in the mid-1980s.

## Technical comments

The unit of the landings are not stated in tables and figures.

## Conclusions

There are no data allowing an assessment of stock trends.

## Cod in Division VIa (West of Scotland) (Report Section 3.2)

1 ) Assessment type: update
2 ) Assessment: analytical
3 ) Forecast: TSA-short term
4 ) Assessment model: TSA
5 ) Consistency: This year's assessment is very similar to the results from last year
6 ) Stock status: SSB 5166 t in 2009 is below $\mathrm{Blim}_{\lim }(14000 \mathrm{t})$ and Z-0.2 (0.89) is estimated to be above $\mathrm{F}_{\lim }(0.8)$ however the Z-0.2 estimate has very large confidence intervals.
7 ) Management Plan: Cod in Division VIa is included in the EU long-term management plan for cod stocks and the fisheries exploiting those stocks (Council Regulation (EC) 1342/2008). The plan and its evaluation by ICES were addressed by WGCSE 2009.

## General comments

The WG addressed the ToRs by providing an update with associated management advice.

Some of the Reviewers found the text hard to follow and considered the Report could be better structured.

The assessment was carried out according to the Stock Annex description. The RG could not run the TSA model and could only check the documented inputs which are in accordance with the annex. The results were carried over correctly to the advice sheets.

The WG Report and Stock Annex only include ecosystem information in respect to the possible role of seal predation as a source of the unaccounted mortality. Climate effects on cod production are not considered.

The stock is subject to an EU multi-annual management plan. ICES WGCSE (2009) reviewed the plan in relation to west of Scotland cod and could not conclude that it was precautionary.

The WG has not used mixed fishery data in providing management advice.
Discarding/high-grading appears to be a major problem with discard rates in 2009 as high as $82 \%$ at age 4 . The 2005 year class has been very heavily discarded.

The main issue with this assessment is the exclusion of all fishery (landings and discards) data other than weights-at-age, from 1995 onwards. Effectively the assessment is a survey-based model calibrated against fishery catch-based population estimates pre-1995. The model gives a clear picture of a major decline in abundance at all ages, with a slight upturn in recent years due to improved recruitment of the 2005 and 2008 year classes. The removals predicted from the survey based Z (minus 0.2 for M ) drift progressively away from the WG figures for fishery catches until 2004-2005, then the difference starts to reduce (see WG Figure 3.2.16 below).


As the WG has indicated more accurate catch reporting since 2006 under the Buyers and Sellers legislation, an additional TSA run was conducted by WGCSE reintroducing the landings and discards-at-age data for 2006 onwards. This approach is adopted as the final run for VIa haddock this year. This gives qualitatively the same picture of stock trends as given by the baseline model (WG Report Figure 3.2.23), and gives "true" F estimates for recent years close to Flim (0.8); i.e. very similar to the Z-0.2 estimates for the most recent years from the baseline model. The arguments put forward by the WG for not adopting this as the final model (which would be consistent with VIa haddock approach adopted by the WG this year) are firstly that the model estimates of removals are scaled downwards to an extent that they are the same as the WG catch data in the mid-1990s, whereas the WG states there is evidence for inaccurate catch reporting at that time. Secondly, the TSA model as presently configured can only handle discards up to age 2 whereas high-grading across a wider age range is presently occurring. The first of these arguments may be spurious, as the confidence intervals around the model predicted removals in the 1990s are wide and could encompass the likelihood of misreporting. The second problem (discard age range) is more of an issue and would require reconfiguring of the TSA. Nonetheless, the exercise is useful in indicating that recent $F$ could be of the same magnitude as the Z-0.2 from the baseline model.

## Technical comments

1 ) The WG should explain the differences between the official landings figures and the WG estimates. For example in 2009, do the WG figures include non-official estimates from France and corrections for area misreporting (e.g. to Faroes or North Sea as mentioned in WG Report)?
2 ) Neither the Stock Annex nor the WG Report gives any indication of the sampling effort for estimating discards by fleet/country (other than for 2009 in WG Report). The methods used to raise discards data are only briefly described in the "uncertainties and bias" section of the WG Report.
3 ) Stock weight and catch weight are the same data. More robust and realistic (for younger ages) stock weights could be obtained from Q1 survey data, for example by modelling age and year effects. Q1 fishery data could be included for fully selected ages.

4 ) The table provided in the Stock Annex gives the available input data but not the used input data. The WG should revise the annex to clearly state the "update" assessment inputs (e.g. age ranges in surveys).
5 ) All input data files have a headline saying "with discard" although this is only true for some of the input files.

## Advice sheet

In the advice sheets it is stated that"Catch (landings + discards) is seven times the reported landings". This statement is conditional on the assumption that the reported landings (and the international raised discard estimates) are now accurate. More careful wording would be appropriate.

## Conclusions

The RG considers the updated TSA assessment to be appropriate for providing management advice for the stock, confirming that SSB is well below Blim and fishing mortality is likely to be well above any Fmsy candidates.

The WG has inserted a long and well covered list of future work before the next Benchmark. The RG supports these proposals, but would include the need to view the VIa stock in the context of the western waters metapopulations, and consider the implications of mixing of populations between neighbouring stock areas in VII and in the North Sea (see VIIa cod benchmark proposals).

## Haddock in Divisions VIIb-k (Report Section 7.4)

1 ) Assessment type: Update
2 ) Assessment: Indicator of trends only.
3 ) Forecast: Short-term forecast presented, not used in advice sheets due to uncertainties in the data and the estimate of the large 2009 year class
4 ) Assessment model: XSA, tuned with two surveys and two commercial fleets
5 ) Consistency: XSA has been performed with the same settings as before. Updated results up to 2008 are similar to last year's run
6 ) Stock status: The state of the stock is not precisely known due to uncertainties in the discards data and there are no accepted biological reference points. SSB is perceived to be increasing. The stock is highly dependent on the incoming recruits. Between 2002-2008 no strong year classes have been observed, whilst the 2009 year class seems to be the highest on record.
7 ) Management Plan: None

## General comments

The WG addressed the ToRs relevant to providing advice through an update assessment.

The assessment was carried out according to the Stock Annex description. The RG found no errors in the implementation of the assessment and forecast, and the assessment results were carried over correctly to the advice sheets. No forecast is given in the advice sheets.

The WG Report and the Stock Annex do not contain any information on ecosystem aspects or environmental drivers.

There is no EU management plan for this stock.
The WG has not used mixed fishery data in providing management advice. The proposed increase in square mesh panels to 120 mm will impact other species such as whiting and all such measures should be viewed in the context of mixed species catches.

The WG has done a good job with the available data. The time-series available for the assessment is short although a longer series of landings data are available. The haddock stock in the neighbouring Irish Sea expanded rapidly in the mid-1990s. It would be of interest if the WG plotted reported haddock landings from VIIb-k over a longer period of years to provide a longer-term picture.

Landings show a marked increase from 2008 to 2009. Recent ICES advice has been for no increase in effort. In general, reported effort statistics show declining effort in the French gadoid fleet (which takes most of the haddock catch) and stable effort in the Irish fleets, so the landings increase represents improved availability of haddock in 2009.

As with other stocks, official French data were not received for 2009 and the WG obtained alternative figures for the assessment (not described). There is also a large discrepancy between Irish official landings data for 2009 and the WG estimates.

The major source of uncertainty in the assessment is the estimates of discards which are based on very small numbers of Irish observer trips, extrapolated to all fleets.

There are also some uncertainties concerning landings data. Discarding appears to be an important feature in the fishery, comprising a large fraction of the catches up to age 3 and including fish above MLS. With this stock and many others in WGCSE, it is disconcerting that after eight years of EU Data Collection Framework requirements to estimate discards (to a given precision level) and transmit the data to ICES WGs, that estimates provided by ICES remain incomplete and of such poor quality, even for heavily discarded species such as haddock that are caught widely in trawl fisheries. The WG should highlight this issue.

Survey data for VIIb-k haddock have variable area coverage but appear to be consistent in tracking year classes, and provide a consistent index showing a potentially very large 2009 year class. The potential of the new Irish Q1 joint science-industry survey to provide additional data on haddock is not mentioned.

## Technical comments

1 ) The data for this stock are patchy and subject to large variability. The RG agrees with the advice from last year's RG that the data for VIIb-k haddock are more suitable for inclusion in a statistical modelling framework in which the nature and magnitude of the errors in the different datasets are accounted for, and the bias and variance in the population estimates can be properly evaluated. This should be explored before any future benchmark.
2 ) The WG has not included F and N at-age tables from the assessment, and does not provide a graphical comparison with last year's assessment.
3 ) The WG has declined to put forward any candidate FmSY values due to concerns that the stock-recruit relationship is not well captured. It is clear that the Ricker model would not be a robust choice. However, the range of $B \& H$ and hockey stick $\mathrm{F}_{\mathrm{msy}}$ and yield-per-recruit $\mathrm{F}_{0.1}, \mathrm{~F}_{35 \%}$ and $\mathrm{F}_{\mathrm{max}}$ values for landings are within the range 0.18-0.26.
4 ) The Stock Annex refers to the use of ALKs derived from all years combined, to estimate age compositions for young haddock where no age data have been collected. Unless it is for the very youngest age classes with effectively no significant overlap in length-at-age, this will smooth out yearclass signals.
5 ) In Summary Table 7.4.9, it would be useful to give the $\mathrm{F}(2-5)$ separated by landings and discards.

## Conclusions

The WG has taken note of earlier recommendations from RG and made a good compilation of the needed steps to perform before making a benchmark assessment. Particular progress is needed on obtaining and making best use of available discards data, and evaluating the quality of such data using the ICES Quality Assurance Framework.

The RG agrees that the WG assessment is suitable for indicating general trends but not for providing forecasts. However, the forecast has a useful exploratory role, given the possibility of a very large 2009 year class entering the fishery. Given a scenario that this is true, the weight discarded would increase from around 7-9 thousand tonnes ( $\pm$ some unknown but very large standard error) to over 20 thousand tonnes in 2010 and 2011. A key question is where in the Celtic Seas this abundance of small haddock will be located, and hence where would measures to avoid discarding be
best targeted. The WG should monitor the distribution of these fish through surveys and observer data and provide managers with this information.

## Haddock in Division VIIa (Irish Sea) (Report Section 6.3)

1 ) Assessment type: update
2 ) Assessment: survey trends
3 ) Forecast: none
4 ) Assessment model: Single fleet SURBA analysis, using only the NIGFSMar survey
5 ) Consistency: Updated survey trends are very consistent with last year's assessment.
6 ) Stock status: uncertain. SSB has decrease since 2008. Recruitment in the last year appears to be above average. Total mortality appears relatively stable.
7 ) Management Plan: There is no specific management plan for haddock in the Irish Sea. Due to the bycatch of cod in the haddock fishery, the regulations affecting Irish Sea haddock remain linked to those implemented under the cod management plan (Council Regulation (EC) 1342/2008).

## General comment

The assessment was carried out according to the Stock Annex description and the WG addressed the ToRs. The RG found no errors in the assessment.

Mixed fishery data are not used when formulating management advice by the WG. The Advice Sheet notes that discarding of small haddock is substantial in the Nephrops fisheries and recommends the use of 120 mm square mesh panels.

The WGCSE was unable to provide absolute values for $\mathrm{F}_{\text {msy }}$ or $\mathrm{F}_{\text {msy }}$ proxies, as there are insufficient data to derive absolute estimates of $\mathrm{F}_{\text {msy }}$ with any degree of precision.

The method used by the WG to estimate stock weights based on survey data could be further developed and applied to other stocks with good survey data but noisy stock weights from fishery sampling at older ages.

The survey data show very coherent year-class signals and appear to give a very clear picture of the development of the stock.

## Technical comments

1 ) The WG states that sample-based evidence suggests that WG estimates are close to reported landings since 2006. This could be made more explicit in the landings tables (e.g. a statement such as "since 2006, officially reported landings are used by the WG as sample-based evidence confirms more accurate catch reporting since 2006").
2 ) The WG should make it clear in the Table 6.33 and 6.34 headings that it is landings-at-age, not catch-at-age.
3 ) In Table 6.3 .4 weights in 2003 are not present; however they are found in the input data for that year. The derivation of the figures in the input file should be explained.
4 ) The WG shows the results of Annual Egg Production Survey estimates of haddock SSB as relative trends confirming the trends in the March GFS between 2006 and 2008 but does not cite the absolute estimates ( 6 kt [CV $32 \%$ ] in 2006 and 9.5 kt [CV 24\%] in 2008). These are very large compared to the WG landings of 650-870 t for these years, even if the discarding-at-
age 2 was taken into account. This would imply a much lower mortality than given by the age profile in the groundfish surveys (which indicate $Z$ of around 1.5). However there is no evidence from any fishery data for an age composition that would reflect low mortality.
5 ) The main problem with the historical yield-per-recruit analysis is the absence of discard fishing mortality.

## Advice sheet

The advice sheet was only partially completed the time of review. The correct plots were carried over from the WG Report.

## Conclusions

The RG considers the updated survey-based analysis to provide an appropriate basis for formulating management advice based on relative abundance trends. The different surveys provide a consistent picture of the stock development. The SSB indices appear to respond dynamically to the very variable recruitment, as would be expected given the steep age profile in the surveys. Stock trends indicate an increase in SSB over the time-series followed by a decrease since 2008 due to some belowaverage year classes. Recruitment in the last year appears to be above average which is expected to halt the decline in SSB. The index of total mortality appears relatively stable.

The state of the stock with regard to reference points is uncertain as there are no biological reference points calculated, and the fishing mortality cannot be estimated directly from the surveys without independent knowledge of the survey selectivity characteristics across the age classes.

Discarding of haddock can be substantial, and the WG should have a priority to derive appropriately raised and quality-assured discards-at-age estimates, following the guidelines in the ICES Workshop on Discard Raising Procedures (which can be implemented through COST tools) and evaluating precision and bias. The present tabulation of historical discards data (Table 6.3.5) is unwieldy and does not give a clear picture of the trends in discarding.

Given the availability of data other than those used in the survey assessment (other survey data; egg production estimates; discards data) there is an urgent need for a data compilation workshop and benchmark assessment for this stock to establish a more comprehensive evidence base and a robust quantitative procedure for developing management advice. Benchmarking alongside the VIIe-k stock would be beneficial.

## Haddock in Division VIb (Rockall) (Report Section 4.3)

1 ) Assessment type: Update
2 ) Assessment: analytical
3 ) Forecast: Short-term forecast provided
4 ) Assessment model: XSA
5 ) Consistency: Updated assessment results in 60\% upward revision of 2008 F and a small downward revision of SSB. Assessment does not exhibit retrospective bias but appears to be unstable due to weak shrinkage used with noisy data. The survey-series has been revised since last year.
6 ) Stock status: $\operatorname{SSB}$ is currently well above $B_{p a}$ and $F$ is below $F_{p a}$ and close to F giving long-term equilibrium yield.
7 ) Management Plan: None

## General comments

The WGCSE addressed the ToRs in providing an updated assessment with associated management advice.

The assessment was carried out according to the Stock Annex description. However the method for deriving the abundance indices for the Rockall survey was altered in 2009 to exclude sporadically sampled strata, resulting in a new tuning file. This appears to have been done by the stock assessor rather than the lab that conducts the survey, and may not be adequately quality-assured. This results in some substantial revision (e.g. halving of the indices in 1992 and 2006) but there is almost no discernible difference to SSB, F and recruitment trends using the two series for an assessment using data up to 2008. The addition of 2009 data however causes the assessment with the "old" survey data file to fail (no convergence) and the new survey data to give a large residual and implausibly low F at age 5 in 2009. During the Advice Drafting Group it also came to light that the Scottish survey data for 2008 had been reworked by the assessor using Russian ALKs, which may further have degraded the consistency of the series. See Technical Comment 2 below.

The RG found no obvious errors in the implementation of the assessment and forecast (although the RG has concerns about the model settings and re-worked surveyseries). No RCT3 input and output tables were provided so these could not be checked. The results were carried over correctly to the advice sheets apart from two errors in the forecast table headers that ICES has been advised of.

Ecosystem aspects: The Stock Annex describes closures on Rockall to protect vulnerable habitats. In order to protect cold-water corals, three areas (North West Rockall, Logachev Mounds and West Rockall Mounds) are closed since January 2007. A new area to protect cold-water corals (Empress of British Banks) was established by the NEAFC in 2007. These are in addition to the Rockall Haddock Box to protect prerecruits that has been in place since 2002.

There is no EU management plan in place for this stock.
The WG has not used mixed fishery data in providing management advice. The Stock Annex and WG Report do not provide information to evaluate bycatches in the directed haddock fishery.

A major signal in the fishery data is the very sharp reduction in Russian landings of haddock from around 5000 t in 2004-2005 to only 55 t in 2009. UK landings declined in the mid-2000s but have subsequently increased again.

Substantial discarding occurs in the EU fisheries but has only been directly estimated in a few years. For other years, the WG conducts a convoluted process to infer discards from the Scottish survey length frequencies together with theoretical selection parameters and a discarding ogive, tuned using data from the few historical years with observer data. It remains unclear if these estimates are robust over the full timeseries since there appears to have been no observer data since 2001 other than a few trips sampled by Ireland in 2007-2009. The WG has not used the latter data to check their consistency with the imputed values. Improved sampling of discards is needed. The WG recommends the need for technical measures to reduce discarding.

## Technical comments

1 ) Table 4.3.14 highlights 1995, 1997, 1999 and 2001 as years with discards estimates calculated directly from observer trips. The Stock Annex also mentions 1998 and 2000 as years with Irish discards estimates (both are years with no survey). Why were these data not used for estimating discards?

2 ) The large negative catchability residual at age 5 using the new survey data series updated to 2009 may originate from the unusually low survey indices for age 5 in 2009. An unusually low 6-gp index is also apparent in 2009 and may be related to generation of a large positive $Q$ residual for age 5 in 2008. A possible problem could be the dominance of 4-year olds in the survey (2005 year class) which could lead to a tendency to allocate 5-year old haddock as 4-year olds where the ageing is not clear, simply because of the large number of 4 -year-olds in the samples. Anomalously low F is generated at age 5 in 2009 and age 4 in 2008 in the final XSA. This age group has a slope of 2.33 (but low R-square). A large $F$ is induced at age 6 in 2009, age 5 in 2008 and age 4 in 2007 (an apparently weak year class). These impact the $F_{b a r} 2-5$ estimates for 2008 and 2009, which are relatively low. The WG should firstly check the accuracy of the survey index values at ages 5 and 6 in 2009 in the input files, and if these are correct, consider how to deal with the tuning problem to smooth the F's in the terminal year. The assessment is clearly unstable (no convergence was possible with the "old" survey series updated to 2009). Given the noise in the data, stronger shrinkage would be beneficial although this may generate retrospective bias in periods of rapid change in F and stock size. The use of the power model for several age classes also increases the number of parameters to be estimated. During the ADG, the XSA was re-run with the original Scottish survey-series (excluding the changes made to the 2008 and 2009 data), and with stronger shrinkage. This also did not converge rapidly, but provided more coherent catchability residuals. The revised assessment will be included in the WGCSE Report.
3 ) The RCT3 input and output tables are missing and should be provided.
4 ) The Fmsy output table for the different stock-recruit options is not given in the WG Report.
5 ) The RG could not evaluate the significance of the results of the StatCam model, or interpret the differences with the XSA, as the method is not explained anywhere.

## Advice sheets

Short-term forecast table: header says 2010 landings are 3.3 kt . Value should be 3.9 kt (forecast needs updating following ADG re-run of assessment). Status quo F is stated as $\mathrm{F}(06-08)$ when it should be $\mathrm{F}(07-09)$. ICES was advised of the errors.

## Conclusions

The RG is prepared to accept the Advice Drafting Group re-run of the WG final assessment as a basis for providing management advice but is concerned about the instability in the assessment due to noisy data, and the potential errors in the discards data due to the method of calculation from survey data. The forecasted landings at status quo F are in line with recent reported landings levels.

The very low 0-gp survey indices for 2008 and 2009 indicate extremely weak year classes, given the generally good correlation between 0-gp indices and the XSA 1-gp estimates for these year classes. These are reflected in the short-term and mediumterm forecasts which indicate a sharp decline in SSB over the next few years, possibly heading for another trough in the SSB series as observed in the late 1990s and early 2000s.

## Haddock in Division Vla (West of Scotland) (Report Section 3.3)

## 1) Assessment type: Update

2) Assessment: Analytical
3) Forecast: Short term
4) Assessment model: TSA, tuned with two surveys
5) Consistency: The current assessment method has been in use since 2006 and is considered consistent. However a major change is introduced this year in incorporating recent landings and discards-at-age into the TSA. The assessment is therefore no longer a simple update. The retrospective bias for SSB, R or F has decreased in the last years, but is considerable e.g. in the period 1995-2005 particularly in estimating F and SSB.
6) Stock status: ICES classifies the stock as being at risk of reduced reproductive capacity. The stock is below $\mathrm{B}_{\mathrm{pa}}$ since 2005 F has decreased in the same time period without any positive response on SSB and is now estimated to be below $\mathrm{F}_{\mathrm{pa}}$. No strong year classes have been recorded since 2000. There is poor relationship between SSB and R.
7) Management Plan: There is a proposed management plan and it is evaluated by ICES as precautionary

## General comments

The WG addressed the ToRs in providing an update assessment with associated management advice although has altered the basis of the assessment this year from the procedure used last year.

The TSA modeling approach was carried out according to the Stock Annex description, although the input data were altered this year to include fishery data for 20062009. The RG could not run the TSA and was only able to check the model settings. The data files were not available on the SharePoint.

Ecosystem consideration is not discussed in the Report or Stock Annex (which only includes some basic haddock biology under the heading "Ecosystem aspects").

The proposed management plan has been evaluated by ICES and "ICES advises that a harvest rule with a target fishing mortality of 0.3 and a TAC constraint of $\pm 15 \%$ is consistent with the precautionary approach."

The WG provides some indication that haddock is taken in mixed fisheries but does not use mixed fishery data in providing advice in support of management.

Two versions of the Report were available on the SharePoint site, one including the assessment results with the 2006-2009 fishery data incorporated, and the other with the data excluded. As the results from the former appear in the draft advice sheets, this version of the Report has been reviewed.

Catch data is only considered reliable in the periods 1978-1994 and 2006-2009. This is a change compared to last year where catch data after 1994 were not used. In the VIa cod assessment the 2006-2009 catch data is not incorporated even though the same arguments are put forward that the accuracy of fishery data have improved since 2006 due to Buyers and Sellers legislation. Discard estimates probably have derived from the same trips. The main argument put forward by the WG for not adopting the
same procedure for both cod and haddock is that the cod assessment only handles discards up to age 2 (high-grading is occurring recently) and that predation by seals is likely to be a source of significant unaccounted mortality.

Given the major change in the assessment procedure, the WG has not provided a detailed comparison of the trends (and error ranges) for the assessments with and without the 2006-2009 data to give a clearer picture of how the advice has changed due to the new approach. The only comparison in the report is the point estimates for 20082009. It is important that the benefits of having more accurate catch data are clearly demonstrated. Some "unallocated" removals remain for the 2006-2009 period, and have to be removed from the catch forecast.

The SSB estimates from years assessment were SSB (2008)=30 436 t just above $\mathrm{B}_{\text {pa, }}$, and SSB (2009)=20 271 t less than Blim. This year's assessment shifts the estimate down for SSB (2008) to 22114 t , that is around Blim and SSB (2009)=16 818 t (the lowest value on record). The estimate for 2010 SSB=13 336 indicates a steady decline of the stock. The changes between this and last year's assessment is not visible in the retrospective analysis due to the change in data inputs this year.

## Technical comments

1 ) Discard-at-ages $9-15+$ is absent for the whole time period and need not be separately presented in the Table 3.3.4

2 ) The WG has replaced the mean weight-at-age 1 in the Irish landings in 2009 by an average value. It is quite possible that only a few fish were recorded at this age in the landing samples and they were at the upper end of the length-at-age distribution. Unless an error is suspected (in which case the data should be reviewed) there is no justification for over-writing mean weights-at-age in national raised data as this will result in an inconsistency between numbers and weights.
3 ) The WG should clarify if down-weighting of individual data points described in the WG Report were the same as in previous assessments; e.g. ScoGFS Q4 2007 (age 2): was this also down-weighted similarly last year, or has the down-weighting been adjusted according to residuals apparent in the updated assessment?
4 ) If there is no apparent stock-recruit relationship, why is recruitment modeled using a Ricker function in TSA, or is the assessment relatively insensitive to this? The MSY section states that there is no ability to distinguish between Ricker, B-H and hockey stick.
5 ) There is no MSY output Table 3.3.18 in the WG Report, therefore the estimates can only be evaluated from the plots. The Fmsy estimates have poor precision yet the WG provides estimates the estimates in the text as if they are endorsed. The upper limit of these is put in the advice sheets. This is inconsistent with other stocks with poorly fitting stock-recruit curves where the WG is not prepared to accept the Fmsy estimates.
6 ) The WG states that the short-term forecast is done using MFDP whereas it appears to be the Marine Lab software.

## Advice sheet

1 ) The advice sheet was only partially completed at the time of review and had no forecast outputs other than the table header, no YPR figures and no plots for the updated assessment.

2 ) The draft forecast table had incorrect headers: total removals in 2010 given as 8340 t ; correct value is 5350 t . ICES was advised.

3 ) The advice sheet shows only the upper limit of the $\mathrm{F}_{\text {mSY }}$ values given in the Report. This is from the smoothed hockey stick model.

## Conclusions

The RG considers the updated assessment with 2006-2009 catch data included provides a better basis for providing management advice than the previous approach excluding all recent fishery data, although some unallocated removals estimates remain. It addresses possible concerns by industry that more accurate catch reporting in recent years is not reflected in the assessment. However, as noted by the WG, the assessment requires benchmarking to validate the approach and explore other assessment approaches and other not included at present.

The RG aggrees with the WG that on the basis of the revised assessment, the stock currently has reduced reproductive capacity ( $\mathrm{SSB}<\mathrm{Blim}_{\mathrm{l}}$ ). However according to the $\mathrm{F}_{\mathrm{pa}}=0.5$ the stock is harvested sustainably. The short-term forecast indicates that the SSB will not rebuild to $\mathrm{B}_{\mathrm{pa}}$ in 2012 even in the absence of fishing.

The suggested $\mathrm{F}_{\text {mSY }}$ range of $0.19-0.35$ seems appropriate for a haddock stock (e.g. with reference to North Sea haddock), but the arguments for accepting these and rejecting $\mathrm{F}_{\text {msy }}$ estimates from equally badly fitting stock-recruit curves for other stocks is inconsistent.

The RG acknowledges the WG statement of a need for a long-term management plan that takes into account the recruitment characteristics of the stock.

## Benchmark suggestions

1 ) Weight-at-age in stock is derived from weight-at-age in catch "in the absence of a sufficiently long time-series of survey-based weight measurements;" however, 25 years seems to be a sufficient long time-series, and stock weights that include data from Q-1 surveys could be incorporated in the next Benchmark.

2 ) There appear to be linkages with the North Sea stock. Any benchmark assessment of North Sea haddock and VIa haddock should be done in the same meeting to allow these connections and their effect on the assessments to be explored.

## Megrim (Lepidorhombus spp) in Subarea VI (West of Scotland and Rockall) and Subarea IV (North Sea) (Report Section 5.3)

1 ) Assessment type: trends
2 ) Assessment: There is no accepted analytical assessment for this stock
3 ) Forecast: not presented
4 ) Assessment model: cpue trends
5 ) Consistency:
6 ) Stock status: unknown
7 ) Management Plan: Category 6-stock no management plan

## General comments

ICES has not conducted an analytical assessment of this stock since 1999. The assessment area has this year increased and includes now also Area IV and IIa. The commercial effort has in recent years decreased in Area VIa, however for 2009 only Irish effort data is available.

Only $53 \%$ of the overall TAC was used. No landings data were made available to the WG by Spain or France therefore the uptake during 2009 will be higher than indicated by the official figures. Historically, France only utilizes $\sim 10 \%$ of its available quota, Spanish uptake has been $\sim 80 \%$.

The survey and fishery catch rates presented in the Report suggest an increase in abundance of megrim around Scotland since the mid-2000s. The Irish lpue data from VIb does not show an increase. There is therefore consistent evidence that megrim populations in VIa and IV are at least stable, and probably increasing, but the picture for VIb is less clear.

## Technical comments

1) Section 5.3.2 Discards data are stated as only available from Ireland, and only three trips were sampled. The year of sampling is not mentioned but is presumed to be 2009 as last year's WG Report indicated nine Irish trips were sampled. Additional information is given from Laurenson and McDonald (2008) but the fleet sampled is not referred to. Sampling and provision of data to the WG will need to be adequate to demonstrate if discard rates decline as expected from the selection curves given in the WG Report, following the increase to 120 mm mesh.
2 ) Section 5.3.2 surveys: The WG Report states that the ratio of 2007-2009 biomass indices to 2003-2006 is 28 and $53 \%$ for ICES Area VI and IV (partial coverage) respectively. According to Table 5.3.4 the biomass increase for VIa is $28 \%$ and the increase for IV is $23 \%$, not $53 \%$.

## Conclusions

The RG considers that the survey and fishery lpue data for VIa and the northern North Sea are consistent with the stock being at least stable and probably increasing since the mid-2000s. This may reflect the large decline in fishing effort. However the population dynamics underlying the trends (recruitment and F) are not known.

The aim is to benchmark VI and IV megrim in 2010/2011. Area misreporting was not taken into account this year, however as data from 1996 and 1997 indicate misreport-
ing in the area of $50 \%$ of estimated landings for Division VIa this should be taken account for in the Benchmark. Exploring other models e.g. SURBA, Catch-Survey Analysis (CSA), etc would be interesting in a benchmarking.

The RG agrees that the Benchmark for megrim should include all the megrim stocks covered by WGCSE and WGHMM.

## Nephrops in Division Vla (FU 11 North Minch) (Report Section 3.5)

1 ) Assessment type: Update
2 ) Assessment: Fishery trends and absolute abundance estimates from UWTV survey .
3 ) Forecast: short-term prediction of landings for 2010 at various harvest ratios using catch option table developed during the Benchmark.

4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG (WKNEPH, 2009).
5 ) Consistency: Methods are the same as last year plus attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES.

6 ) Stock status: The UWTV series indicates that abundance has declined from the high estimates in 2003-2006 to around the same values in 2000-2002. The current harvest ratio is above the $\mathrm{F}_{\text {msy }}$ proxy. The length-frequency distribution, mean size and mean weight of Nephrops have all been stable for the time-series. Since current harvest rate is above the $\mathrm{F}_{\text {msy }}$ proxy, the transition scheme towards the ICES MSY framework applies.
7 ) Management Plan: There is no management plan for this stock.

## General comments

The assessment was carried out according to the Stock Annex description and the WG addressed the ToRs. Errors were found in two key tables of historical weights and harvest ratios (see technical comment 2) but the catch forecasts appear to have used the correct values.

The RG was not able to run the mentioned "Bell/Dobby combined sex-length cohort analysis (LCA) model". So, the calculation of the Fmsy proxy candidates could not be reviewed.

This year's assessment is performed using combined length compositions from trawl and (new) creels. The WG considers the incorporation of creel length compositions has improved the estimates of harvest ratios. However the effect on $\mathrm{F}_{\mathrm{msy}}$ estimates and catch forecasts of including these data is not explored.

Length compositions and mean weights have been relatively stable over time.
Discards are included in the assessment and forecast. The WG states that discards are sampled adequately for the fishery, although no statistics are provided to indicate bias and precision.

## Technical comments

1) In the WG Report Section 3.5 it reads "The stock is being exploited unsustainably" the same section also quotes ICES advice for the same year "The current fishery appears sustainable".
2 ) Table 3.5.9 (mean weights in landings, FU11-13) and Table 3.5.10 appear to be wrong. The mean weights (07-09) do not conform to the values used in the catch forecasts. The harvest ratios in Table 3.5.10 appear to be landings numbers (not total catch numbers) divided by adjusted survey numbers and are different from the values plotted in the advice sheets. This was
checked at the ADG and both the harvest ratios and the discard rates appear to be incorrect for FU11. Corrected Tables 3.5.9 and 3.5.10 need to be inserted. Table 3.5.10 should in any case be replaced with a table similar to Table 6.5.7 in the FU15 report, laying out all the key variables.
3 ) In Section 3.5.7 the text reads; "Table 3.5.10 also shows the estimated harvest ratios over this period. These range from $7-32 \%$." But they range from $13 \%$ to $34 \%$. The $7 \%$ is a discard rate. This text is incorrect anyway as the Table has the wrong harvest ratios.
4 ) Sampling levels for discards are given, but no discards estimates are provided other than discard rates.
5 ) Legend for Table 3.5 .8 should indicate the burrow counts are not biasadjusted.
6 ) The "Conclusions of the Review of the 2009 assessment" cited in the 2010 Report are not the ones for FU11 but the ones for FU13 (Clyde).
7 ) The WG presented some ecosystem aspects in Section 3.5.1 and referred to further information in the Annex, although this section is empty in the Annex.
8 ) The WG states that the UWTV bias factors include expert judgment, but there is no knowledge of the precision of this, or of the constancy of the bias factors (e.g. edge effects) when burrow density is changing. This needs further investigation.
9 ) As with other Nephrops stocks, more accurate catch reporting since the introduction of Buyers and Sellers legislation has resulted in an apparent increase in landings and lpue which causes misleading trends plots (Figure 3.5.4). This should be clearly indicated on the figure legends.

## Conclusions

The RG considers the Underwater Television Survey (UWTV) and associated catch options to be an appropriate basis for management advice.

The RG agrees with the WG that management of this stock should be applied at a local FU level rather than at the ICES Division level.

The RG agrees with the WG that $\mathrm{F}_{35 \%}$ spr (combined between sexes) is consistent with the approach adopted by the WGCSE for choosing $\mathrm{F}_{\text {msy }}$ proxies for Nephrops.

If ICES is to use UWTV abundance estimates as absolute, then biases due to incomplete coverage of Nephrops habitat need to be evaluated. The RG agrees that the relationship between fishing area (VMS) and survey area need further exploration.

## Nephrops in Division Vla (FU 12 South Minch) (Report Section 3.6)

1 ) Assessment type: Update
2 ) Assessment: Fishery trends, and UWTV survey estimates
3 ) Forecast: Short-term prediction of landings for 2010 at various harvest ratios using catch option table developed during the Benchmark

4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG (WKNEPH, 2009).
5 ) Consistency: Methods are the same as last year plus attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES.

6 ) Stock status: The UWTV series indicates that abundance has declined from the high estimates in the early 2000s to around the same values in the late 1990s. The harvest ratio in 2009 was close to the $\mathrm{F}_{35 \%}$ spr $\mathrm{F}_{\text {msy }}$ proxy. The length-frequency distribution and mean size of Nephrops have all been stable for the time-series.

7 ) Management Plan: None

## General comments

The assessment and provision of advice in 2010 followed the process defined by the Benchmark WG and the WG fulfilled the ToRs.

At the time of initial reviewing the advice sheet was not completed.
Ecosystem consideration is briefly described in the WG Report and refers to more information in Stock Annex, where this section is empty.

The RG was not able to run the mentioned "Bell/Dobby combined sex-length cohort analysis (LCA) model", so the calculation of the FmSy proxy candidates could not be reviewed.

This year's assessment is performed using combined length compositions from trawl and (new) creels. The WG considers the incorporation of creel length compositions has improved the estimates of harvest ratios. However the effect on Fmsy estimates and catch forecasts of including these data is not explored.

Length compositions and mean weights have been relatively stable over time.
Estimates of discard rates are included in the assessment.
The WG gives a good overview of the MSY work done and clear view of the preferred $\mathrm{F}_{\mathrm{msy}}$.

The technical aspects and general fishery information is well described in the Stock Annex.

## Technical comments

1 ) The catch forecast table in Section 3.6 .8 gives the wrong mean weight in the landings. This is given as 21.3 g whereas the value used in the forecast is 23.8 g . Figure 21.3 links to the incorrect Table 3.5.9 in the FU11 Report.

2 ) The RG recommends replacing Table 3.6 .5 with a table similar to Table 6.5.7 in the FU15 Report, laying out all the key variables.

3 ) The discard rates in Table 3.6.5 (if correct) indicate that the $16.7 \%$ rate used in the forecast could be slightly below recent averages.
4 ) No discards estimates are provided; just discard rates.
5 ) Legend for Table 3.6.4 should indicate the burrow counts are not biasadjusted.
6 ) The WG states that the UWTV bias factors include expert judgment, but there is no knowledge of the precision of this, or of the constancy of the bias factors (e.g. edge effects) when burrow density is changing. This needs further investigation.
7 ) As with other Nephrops stocks, more accurate catch reporting since the introduction of Buyers and Sellers legislation has resulted in an apparent increase in landings and lpue which causes misleading trends plots (Figures 3.6.1 and 3.6.2). This should be clearly indicated on the figure legends.

## Conclusions

The RG considers the Underwater Television Survey (UWTV) and associated catch options to be an appropriate basis for management advice.

The RG agrees with the WG that management of this stock should be applied at a local FU level rather than at the ICES Division level.

The RG agrees that $\mathrm{F}_{35 \% \text { spr }}$ (combined between sexes) is consistent with the approach adopted by WGCSE for choosing $\mathrm{F}_{\mathrm{msy}}$ proxies for Nephrops.

If ICES is to use UWTV abundance estimates as absolute, then biases due to incomplete coverage of Nephrops habitat need to be evaluated. The RG agrees that the relationship between fishing area (VMS) and survey area need further exploration.

An improvement suggested by WG and endorsed by RG is improving the coverage and timing of the UWTV survey and correlating it with VMS data for best adjustment to the harvest area.

## Nephrops in Division VIa (FU 13 Firth of Clyde) (Report Section 3.7)

1 ) Assessment type: update
2 ) Assessment: Fishery lpue trends and UWTV survey estimates
3 ) Forecast: Short-term prediction of landings for 2010 at various harvest ratios using catch option table developed during the Benchmark.

4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG (WKNEPH, 2009).
5 ) Consistency: Methods are the same as last year plus attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES.
6 ) Stock status: As with the other VIa FUs. The UWTV series indicate that abundance has declined over the last three years, in this case following a progressive increase from low values in the mid 1990s. Recent harvest ratios are above the $\mathrm{F}_{35 \% \mathrm{spr}} \mathrm{F}_{\text {msy }}$ proxy. The length-frequency distribution and mean size of Nephrops have all been stable for the time-series. The Sound of Jura Subarea appears to have very low harvest ratios.

7 ) Management Plan: none.

## General comments

The assessment and provision of advice in 2010 followed the process defined by the Benchmark WG and the WG fulfilled the ToRs.

For the first time an attempt is also made to use the UWTV data available for the Sound of Jura Subarea. Although the dataseries is incomplete it indicates a lower burrow density and lower harvest ratios than in the Clyde.

As with other Nephrops stocks, more accurate catch reporting since the introduction of Buyers and Sellers legislation has resulted in an apparent increase in landings and lpue which causes misleading trends plots (Figure 3.7.1 and 3.7.2). This should be clearly indicated on the figure legends.

The length composition indicators ( $>35 \mathrm{~mm}$ ) are relatively stable over time. A good upcoming recruitment may be indicated as the means size of females ( $<35 \mathrm{~mm}$ ) in the catches are at the lowest observed boundaries since two years.

Estimates of discard rates are included in the assessment.

## Technical comments

1) The combination of a sharp dip in burrow count and increased landings in 2007 gives a very large harvest ratio ( $\sim 0.5$ ) yet the burrow count the following year increases to a value more in keeping with the general trend. This would be unexpected if such a large fraction of the stock had been caught, unless the burrow density is heavily driven by new recruits.

2 ) The RG recommends replacing Table 3.7.8 with a table similar to Table 6.5.7 in the FU15 Report, laying out all the key variables.

3 ) The discard rate adjusted to account for some survival was estimated at the Benchmark Workshop to be $18.6 \%$ (taking a three year average 20052007) and according to the agreed benchmark protocol this value is used in the provision of landings options for 2011. However the discard rate in 2009 was $39 \%$ and the time-series average is $31 \%$. The landings forecast is
therefore likely to be overoptimistic. The WKNEPH 2009 Report suggests keeping the discard rate in forecasts at the value estimated by WKNEPH. However, the mean value of $18.6 \%$ for 2005-2007 is very different from the mean for those years in Table 3.7 .8 ( $\sim 33 \%)$. This was raised at the ADG but not adequately resolved.
4 ) In Section 3.7.7, first paragraph, the Fmsy harvest ratio for the Clyde is given as $13 \%$ whereas the stated forecast landings of 3558 t is for the proposed $\mathrm{F}_{\text {max }}$ ratio of $16.5 \%$.
5 ) The WG states that the effort and lpue are not reported due to the problem with a discontinuity around 1995, however the data are presented in Figure 3.7.2. It would be useful to have a consistent effort series. The WG should indicate clearly on the plot legends that the landings and lpue data from ~2006 onwards are likely to have increased due to more accurate catch reporting.

## Conclusions

The RG considers the Underwater Television Survey (UWTV) and associated catch options to be an appropriate basis for management advice, but is concerned about the possible overestimate of landings in the forecast due to the use of a discard rate well below the recent average.

The RG agrees with the WG that management of this stock should be applied at a local FU level rather than at the ICES Division level.

The RG agrees that $\mathrm{F}_{\max }$ (harvest ratio $16.5 \%$ combined between sexes) is consistent with the approach adopted by WGCSE for choosing $\mathrm{F}_{\mathrm{msy}}$ proxies for Nephrops. This is predicted to deliver an $\mathrm{F}_{35 \% \mathrm{spr}}$ of about $20 \%$ for males. The use of the low UWTV estimates from the mid-1990s to give a $B_{\text {trigger }}$ of 579 million individuals is appropriate as a first estimate but has no basis other than being a low point in a relatively short-time series.

The WG was not able to conduct a yield-per-recruit for the Sound of Jura population and has adopted the Clyde Fmsy calculations as an interim approach (combined sex $\mathrm{F}_{35 \% \mathrm{SpR}}$ HR of $13 \%$, based on low burrow density). The RG notes that the discard rates appear to be negligible which means that the Fmsy estimates for the Clyde (where an $18.6 \%$ discard rate was adopted) may have an additional bias. The Btrigger point for this FU (bias adjusted lowest observed UWTV abundance) has not been defined but is expected to be below 200 million individuals. RG agrees on this provisional figure as the approach is consistent with the other VIa FUs.

## Nephrops in Division VIIa (FU 14 Irish Sea East) (Report Section 6.4)

1 ) Assessment type: Update
2 ) Assessment: Fishery lpue trends and UWTV survey estimates
3 ) Assessment: Forecast: presented
4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WG (WKNEPH, 2009).
5 ) Consistency: New approach and advice based on UWTV survey
6 ) Stock status: provisional estimate
7 ) Management Plan: None.

## General comments

The assessment and provision of advice in 2010 followed the process defined by the Benchmark WG and the WG fulfilled the ToRs.

The assessment determines the health of the stock by looking at trends in total landings, lpue, size composition, and biological data from the commercial fisheries. For the first time for this stock the results from UWTV survey data were used to calculate provisionally absolute abundance estimates for 2009.

Presentation was clear with respect to updated data tables and figures.
The RG was not able to run the mentioned "Bell/Dobby combined sex-length cohort analysis (LCA) model". So, the calculation of the Fmsy proxy candidates could not be reviewed.

## Technical comments

1 ) As with other Nephrops stocks, more accurate catch reporting since the introduction of Buyers and Sellers legislation has resulted in an apparent increase in landings and lpue which causes misleading trends plots (Figure 6.4.1-6.4.3) This should be clearly indicated on the figure legends.

2 ) The RG recommends including a table similar to Table 6.5.7 in the FU15 Report, laying out all the key variables.
3 ) The WG does not provide any discard estimates other than what can be inferred from Figure 6.44, and should provide a table. It is not possible to evaluate the quality of the data from what has been presented, and it is noted that the discard sample rates in 2009 indicate a substantial decline in sampling in 2009. With no knowledge of the variance of the discards estimates, the significance of the apparent very sharp drop in discarding in 2009 cannot be evaluated.

4 ) The forecast table in Section 6.4.5 gives total numbers caught under the header "retained number."

## Conclusions

The RG agrees that the UWTV survey and associated $\mathrm{F}_{\mathrm{MSY}}$ values represent an appropriate means of providing quantitative management advice, but notes the short timeseries and the absence of stock-specific growth rates which are inferred from FU15.

The indications of a recent decline in recruitment should be revisited in 2011 and if the 2010 UWTV is successful there opens the opportunity to assess this stock more reliably on an annual basis.

RG also again agrees with WG that TAC allocations should be made at an FU level instead of a single TAC for the entire Area VII in order to prevent localized overfishing.

## Nephrops in Division VIIa (Irish Sea West) (Report Section 6.5)

1 ) Assessment type: Update
2 ) Assessment: Fishery lpue trends and UWTV survey estimates
3 ) Forecast: Landing predictions for 2011 presented
4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WK (WKNEPH, 2009).
5 ) Consistency: Methods are the same as last year plus attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES.
6 ) Stock status: The UWTV survey together with trawl survey data gives the impression of a constant high abundance over the time-series. Reported landings have been stable around 9000 tonnes without a negative impact on the stock.
7 ) Management Plan: None.

## General comments

The assessment was carried out according to the Stock Annex description and the WG addressed the ToRs. The RG found no errors in the assessment.

The RG was not able to run the mentioned "Bell/Dobby combined sex-length cohort analysis (LCA) model". So, the calculation of the FmSY proxy candidates could not be reviewed.

The fishery is well described in WG and Stock Annex. Mixed fisheries data are not used in providing management advice. Discards estimates are included in the assessment but not tabulated in the Report as weights or numbers, only as proportions.

The WG Report is succinct but could be structured more clearly. The WG should establish what types of information should go into different sections. For example the Section "Historical stock development" is more about the UWTV survey design and biases than about historical stock development. Ecosystem considerations are not described in the appropriate section in the WG Report, but information occurs in other sections including in the MSY explorations.

There is no management plan for this stock but the fishery is affected by measures implemented for cod. The cod closure affects the distribution of fishing to some extent, and the types of gears used (e.g. Swedish grids being used by some vessels). The effort control regime has also influenced the switching of effort into the Nephrops fishery.

## Technical comments

1 ) Table 6.5.6: the header does not make it clear if the $10 \%$ survival of discards is included in the figures in the table or applied afterwards. The figure of $20 \%$ for 2008 appears low compared to surrounding values; the WG should provide information on discard sampling rates for the different fleets, and ideally precision and bias evaluations, considering the sensitivity of the landings forecasts to assumed discard rates.
2 ) As with other Nephrops stocks, more accurate catch reporting since the introduction of Buyers and Sellers legislation has resulted in an apparent in-
crease in landings and lpue which causes misleading trends plots (Figure 6.5.1) This should be clearly indicated on the figure legends.

3 ) The scaling of the UWTV estimates vs. the Trawl survey (Figure 6.5.5) is heavily influenced by the 2003 and 2004 trawl estimates which look like possible year effects in the survey. If the two series were rescaled so that the UWTV and trawl survey data for 2005 onwards lie on top of each other, the 2003 and 2004 UWTV figures would probably be more in accordance with the trawl data up to 2002. This would bring the UWTV estimates closer to a re-calculated $\mathrm{B}_{\text {trigger }}$ level. The WG should evaluate the 2003 and 2004 trawl survey data to see if there is an explanation for the high values. The high values in 2003 and 2004 are not apparent in the April survey (Figure 6.5.2).
4 ) The WG should provide the time-series of fishery landings-discards length compositions given for other FUs in VIa and VII, to allow a more direct comparison.

## Conclusions

The RG agrees that the UWTV survey and associated $\mathrm{F}_{\mathrm{MSY}}$ values represent an appropriate means of providing quantitative management advice. However there are concerns about the inclusion of a very low discard rate estimate for 2008 in a 2 -yearmean value used in the forecast.

The RG agrees with the WG that management on a FU level would be beneficial.
The FU15 stock is the most abundant and most densely packed of the assessed Nephrops stocks. Official landings fluctuated around $\sim 9000$ tonnes over a period in the 1990s during which the stock increased according to the trawl surveys. The more recent abundance indices from UWTV surveys and trawl survey data in the 2000s (excluding 2003 and 2004 August surveys) have remained stable whilst the stock has yielded landings of around 9000 t .

The RG agrees that $\mathrm{F}_{\max }$ (harvest ratio $17.1 \%$ combined between sexes) is consistent with the approach adopted by WGCSE for choosing $\mathrm{Fmsy}^{\text {msy }}$ proxies for Nephrops. This is predicted to deliver an SPR for males of $28 \%$ virgin SPR. The RG considers that the method adopted for estimating a $B_{\text {trigger }}$ is strongly influenced by possible year effects in the trawl survey in 2003-2004, and that the rescaling of the trawl survey estimates would be better done using 2005-2009 data.

The bias correction factor needs further investigation including, as suggested by the WG a precision estimate.

## Nephrops in Division VIIb,c,j,k (FU 16 Porcupine Bank) (Report Section 7.6)

1 ) Assessment type: update
2 ) Assessment: trends
3 ) Forecast: not presented
4 ) Assessment model: No Analytical Assessment
5 ) Consistency: Consistent with last assessment
6 ) Stock status: Status of the stock cannot be evaluated because reference points have not been determined for this stock, although the stock is perceived to be over exploited and on the point of collapse.
7 ) Management Plan: There is currently no management plan for this stock, but there are area closures, MLS and mesh size regulations.

## General comments

The assessment was carried out according to the Stock Annex description and the WG addressed the ToRs in providing updated series of indicators.

General ecosystem information has not been provided, and mixed fishery data are not used in support of management advice.

The main indicators are poor recruitment in 2004-2008 based on \% of LFD $<32 \mathrm{~mm}$ CL, with indications of a return to more "normal" level in 2009. A proxy for Z based on the LFDs shows increasing $Z$ in the 2000s, and a very large reduction in \% males in the landings and survey has been apparent in the last 2-3 years. Discards data are not used in this assessment, and discard levels are unknown.

## Technical comments

1 ) It is difficult to compare the effort and lpue series for the Irish, French and Spanish fleets (Figure 7.6.9) as the Spanish data include an engine power correction and the others are in "hours trawling". It is also difficult to see why the lpue (kg per h trawling) should be so different in the French and Irish fleets unless there are differences in the accuracy/completeness of landings or effort reporting. The lpue series from all countries would benefit from being standardized in the same way.
2 ) It would be useful to mark the area of the closure on the survey maps in Figure 7.6.7, particularly as Figure 7.6.10 has no latitudes or longitudes.

## Conclusions

The RG considers that the indicators of the state of the stock continue to suggest a depleted population with an unnatural sex ratio that could impair productivity. Nonetheless, despite the low proportion male, the stock appears to have produced a better recruitment in 2009, which will need to be protected to promote rebuilding of the biomass. Unfortunately the discard practices are poorly known and sampled and unless this is rectified it will not be known if the 2009 recruitment is being heavily fished and discarded.

The area closure in 2010 may not be fully effective as there appears to be high trawling effort south of the northern "hook", and (in previous years) the time period of the closure covered ca. $60 \%$ of landings (leaving enough time to increase targeting). From
the maps presented it is not clear whether the zones of high recruitment in 2009 are protected in 2010.

## Nephrops in Division VIIb (Aran Grounds, FU 17) (Report Section 7.5)

1 ) Assessment type: Update
2 ) Assessment: Trends
3 ) Forecast: Short term
4 ) Assessment model: Assessment and provision of advice through the use of the UWTV survey data and other commercial fishery data follows the process defined by the Benchmark WK (WKNEPH, 2009).
5 ) Consistency: Methods are the same as last year plus attempts to incorporate decisions taken at WKFRAME for the provision of MSY advice by ICES.
6 ) Stock status: Unclear. The UWTV survey gives a fluctuating abuandance that does not follow the landings. This may be caused by unallocated landings and/or unaccounted natural mortality.
7 ) Management Plan: none.

## General comments

The assessment was carried out in accordance with the description in the Stock Annex and the RG found no errors in the assessment.

At the time of review the advice sheet was not entirely finished. Some figures as basis for the outlook table were missing and the Section "Additional consideration" looked like last year's text.

The assessment approach used by WGCSE 2010 was said to be consistent with that set out in the Stock Annex and WKNEPH (2009). Exploratory SCAs (Separable cohort analysis) were carried out to derive suitable reference points for this stock. The RG could not evaluate the SCA as the input files were not available.

The Stock Annex was very clear and contained good information on ecosystem consideration.

Discard estimates are included in the assessment since 2001 with the exception of 2006-2007 when there was no sampling of landings and discards.

## Technical comments

1 ) During the ADG , feedback from the RG led to an investigation of the inputs to the FMSY calculations (see point (2) below). This led to a discovery of a mistake in the calculations and a revision of the FMSY harvest ratios. The amendments are given below:

| Ref point | Sex | WG harvest ratio | Revised ratio |
| :--- | :--- | :---: | :---: |
| F0.1 | Combined | $6.0 \%$ | $7.2 \%$ |
| F0.1 | Female | $9.1 \%$ | $9.1 \%$ |
| F0.1 | Male | $5.2 \%$ | $6.4 \%$ |
| F35\% | Combined | $9.7 \%$ | $10.5 \%$ |
| F35\% | Female | $13.1 \%$ | $12.8 \%$ |
| F35\% | Male | $6.9 \%$ | $8.4 \%$ |
| Fmax | Combined | $9.7 \%$ | $11.1 \%$ |
| Fmax | Female | $13.2 \%$ | $13.0 \%$ |
| Fmax | Male | $8.1 \%$ | $9.8 \%$ |

The revised forecast table is given below:

|  | Implied fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Harvest rate | Survey Index (millions) | Retained number (millions) | Landings (tonnes) |
| MSY framework | 10.5\% | 552 | 39 | 948 |
| F2010(avg.2007-2009)2009) | 10.6\% | 552 | 39 | 957 |
| F0.1 Combined | 7.2\% | 552 | 26 | 650 |
| Fmax Combined | 11.1\% | 552 | 41 | 1002 |
| F0.1 Comb | 7.2\% | 552 | 26 | 650 |
| F0.1 Female | 9.1\% | 552 | 33 | 822 |
| F0.1 Male | 6.4\% | 552 | 24 | 578 |
| F35\% Comb | 10.5\% | 552 | 39 | 948 |
| F35\% Female | 12.8\% | 552 | 47 | 1156 |
| F35\% Male | 8.4\% | 552 | 31 | 759 |
| Fmax Comb | 11.1\% | 552 | 41 | 1002 |
| Fmax Female | 13.0\% | 552 | 48 | 1174 |
| Fmax Male | 9.8\% | 552 | 36 | 885 |
|  | 2.0\% | 552 | 7 | 181 |
|  | 4.0\% | 552 | 15 | 361 |
|  | 6.0\% | 552 | 22 | 542 |
|  | 8.0\% | 552 | 29 | 722 |
|  | 10.0\% | 552 | 37 | 903 |
|  | 12.0\% | 552 | 44 | 1084 |
|  | 14.0\% | 552 | 51 | 1264 |
|  |  |  |  | Basis |
| Landings Mean Weight (Kg) |  | 0.0 |  | Sampling 2008 and 2009 |
| Survey Overestimate Bias |  | 1.30 |  | WKNEPH 2009 |
| Survey Numbers (Millions) |  | 718 |  | UWTV Survey 2009 |
| Prop. Retained by the Fishery |  | 0.67 |  | Sampling 2008 and 2009 |

2 ) The combined-sex Fmsy proxy harvest-ratios for the Nephrops stocks in VIa and VII other than FU17, all tend to be very similar despite the variations in growth rates and discard rates (see table below). The much lower value for FU17 appears to be due to a low value for females (similar to FU15) and a very low value for males ( $50 \%$ lower than FU15). The same growth data are used for FU15 and FU17. The RG asked for the FU17 model inputs to be checked as the Linfinity for mature females in the Stock Annex table is given as 50 mm but is claimed to be derived from FU15 and FU16 values which are 56-60 mm. The RG was advised that the LCA was run using the same parameters as for the Irish Sea. A source of the large difference between FU17 and other FUs could therefore be a very different length composition and selectivity pattern for males in the 2008-2009 FU17 data than is obtained for the other stocks. The WG should further explore the reasons for the different Fmsy values in FU17, including the quality of the LFDs for landings and discards and the effect of the shift in timing of the fishery in recent years.

| Harvest ratios for different (combined sex) FMSY proxies |  |  |  | Harvest ratios for F35\%spr for males and females |  | Males and imm. females |  | Mature females |  | Burrow densities (per m²) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FU | F0.1 | Fmax | F35\%spr | Male | Female | Linf | K | Linf | K |  |
| 11 | 9.8 | 16.9 | 13.3 | 10.5 | 19.2 | 70 | 0.16 | 60 | 0.06 | 0.55 |
| 12 | 9.7 | 16.9 | 13.1 | 9.8 | 21.1 | 66 | 0.16 | 59 | 0.06 | 0.43 |
| 13 | 9.3 | 16.9 | 13.1 | 9.7 | 22.2 | 73 | 0.16 | 60 | 0.06 | 0.8-1.0 |
| 14 | 9.8 | 16.4 | 13.0 | 14.1 | 12.7 | Same | FU15 |  |  | 0.25-0.38 |
| 15 | 10.6 | 17.1 | 13.4 | 12.5 | 13.5 | 60 | 0.16 | 56 | 0.10 | $\sim 1.0$ |
| 16 |  |  |  |  |  | 75 | 0.14 | 60 | 0.16 |  |
| 17 | 7.2 | 11.1 | 10.5 | 8.4 | 12.8 | 60 | 0.15 | 56* | 0.1 | 0.6-1.4 |
| 20-22 |  |  |  |  |  | 68 | 0.17 | 49 | 0.10 | 0.23-0.4 |

3 ) The Stock Annex and WKNEPH 2009 should be amended to show the correct Linf of mature females in FU17 ( 56 mm if derived from FU15). The RG/ADG was advised that the FMSY calculations were done for FU17 using "the same growth inputs as for the Irish Sea"
4 ) The UWTV estimates and the annual fishery lpue appear to be negatively correlated, although Figure 7.5.7c indicates this may depend in which season lpue is calculated in. If the large UWTV estimate for 2004 was due to strong recruitment, this could generate a subsequent increase in fishery lpue as observed. The fishery LFDs for females show no evidence for a strong recruitment pulse in 2004, although the ones for males shift slightly towards the smaller length classes in 2004 then increase again in subsequent years.
5 ) The Figure captions for Figure 7.59 need more detail to explain what the lines actually are.
6 ) Table 7.5.6 provided clear information on all relevant variables allowing the derivation of historical harvest ratios and forecast inputs to be checked. This should be done for other stocks (a similar table is provided for FU15).

## Conclusions

The RG agrees that the UWTV survey and associated $\mathrm{F}_{\text {MSY }}$ values represent an appropriate means of providing quantitative management advice. The UWTV is a method susceptible to bias but the WG concludes that the survey estimates are considered fairly precise. The RG agrees that $\mathrm{F}_{35 \% \mathrm{spr}} / \mathrm{F}_{\max }$ (both giving harvest ratio of $9.7 \%$ combined between sexes) is consistent with the approach adopted by WGCSE for choosing $\mathrm{F}_{\text {msy }}$ proxies for Nephrops. This is predicted to deliver an SPR for males of $23 \%$ virgin SPR. However the RG still has concerns about the different harvest ratios for males compared to other stocks which should be investigated further. The mean weight in landings and discard rates should also be examined further as they are also key sources of uncertainty.

The bias correction factor needs further investigation including, as suggested by the WG, a precision estimate.

The RG agrees with the WG that management on a FU level would be beneficial.

## Nephrops in Division VIIa,g,j (South East and West of IRL, FU 19) (Report Section 7.8)

1) Assessment type: None

2 ) Assessment: None
3 ) Forecast: None
4 ) Assessment model: -
5 ) Consistency: Cannot be evaluated
6 ) Stock status: The status of the stock cannot be evaluated. Sampling indicates a decline in mean size of indviduals.
7 ) Management Plan: None

## General comments

The FU was not assessed and no data analysis was carried out in 2009 or 2010. It's stated in the ToRs for 2010 that an assessment is to be performed but this is probably an error, there is no history of assessment of this FU.

Advice is given biannually and should be updated in 2010. The advice sheets were not completed at the time of review by one of the Reviewers (2 June) and could not be checked.

The only available information for the FU19 is from the UK March groundfish survey that indicated some decline in mean size. In 2006 there was some UWTV stations covered in the FU19 but there does not seem to be an annual coverage of the area. The WG states that the area and FU is heterogeneous and a UWTV survey abundance is hard to obtain on an accurate and regular level.

No ecosystem considerations are presented in Report.
Discard is not discussed in the text but should be included under the Section named "Commercial Catches and Discards".

The biological sampling could be better described. The sampling is obviously irregular and inconsistent and the time-series is too short to give any information. In 2001 there was a change in sampling methodology and it seems that discard data are included from this point but it is not described.

The fishery description contains information about mixed fisheries and indicates that it is of importance, but there are no indications how it can be used in future advice.

## Technical comments

1) Table 7.8.1 presents landings from FU18 but it is not referred to in text how FU18 and 19 are connected.

## Conclusions

RG agrees that analytical assessment is not possible to perform on this FU
It is recommended that the WG gives some suggestion how the sampling and survey data can be improved.

## Nephrops in Divisions VIIfgh (Celtic Sea, FU 20-22 (Report Section 7.7)

1 ) Assessment type: update
2 ) Assessment: trends
3 ) Forecast: no
4 ) Assessment model: no
5 ) Consistency: consistent with last year's methods and stock annex
6 ) Stock status: unknown
7 ) Management Plan: no

## General comments

The WG has addressed the ToR by providing an update of fishery and survey trends.
No French lpue and effort indices for 2009 were available.
The WG Report and the Stock Annex for this FU are comprehensive but difficult to follow due to the amount of detail.

## Technical comments

1 ) Data from the EVHOE survey were cited but time-series plots of number caught by year, etc, were not found.
2 ) For consistency with other FUs, it would be useful to plot the landingsdiscards length compositions in a column with a line through the mean lengths. Otherwise some of the LFD plots are just duplicating tables with the same data and are hard to interpret.

3 ) Table 7.7.16, showing landings by rectangle and year might be better presented as a series of maps?
4 ) Some values of sex ratio are given in the text, but no data on trends in sex ratio are provided, as for other stocks.
5 ) Figure 7.7.8 gives a "domain area" as in $\mathrm{m}^{2}$. Should this not be $\mathrm{km}^{2}$ to give abundances in millions?

## Conclusions

The RG considers that the indicators provide relatively little information for evaluating stock trends. Irish and French fleets show different lpue trends but it is not clear if this reflects shifts in activities between different grounds.

The UWTV results for the Smalls grounds indicate a stable abundance over the last 34 years. The RG recommends that VMS and other data be used to map out the Nephrops grounds more accurately (see VIa stocks) to allow the possibility of extending the UWTV coverage to include other significant mud patches (Labadie/Nymphe/Seven heads ground) on an annual basis if funding is available for this. The Annex indicates that UWTV was tried on very small areas of these three grounds in 2006 but that poor weather precluded surveys in 2007 and 2008. The different allocation of French and Irish Effort between the four main grounds would argue for UWTV coverage of all areas given the different trends in effort of these fleets. The WG proposal to develop fishery data (length compositions, discard rates, etc.) specifically for the Smalls is a necessity for developing the UWTV survey for providing quantitative management advice for this ground. However the other mud patches should not be ignored.

The back calculation of discard rates when fishing procedures (tailing) are changing is rather difficult and should be replaced by observer data whenever possible. However, the method should be further evaluated for the next Benchmark.

## Plaice in Division VIIb,c (West of Ireland) (Report Section 7.9)

1) Assessment type: No assessment
2) Assessment: None
3) Forecast: None
4) Assessment model: None
5) Consistency: -
6) Stock status: -
7) Management Plan: None.

## General comments

Only landings are presented for this stock. It shows a decrease in landings the last ten years. The WG does not suggest any explanation for this.

## Technical comments

The unit of the landings in Table 7.9.1 is not defined.

## Plaice in Divisions VIIh,k (Southwest of Ireland) (Report Section 7.11)

1 ) Assessment type: SALY
2 ) Assessment: trends
3 ) Forecast: none
4 ) Assessment model: catch-curve analysis and yield-per-recruit analysis
5 ) Consistency: Same approach as last year
6 ) Stock status: Unknown
7 ) Management Plan: None

## General comments

The WG addressed the ToRs by updating the tables and catch-curve analysis.
This is a very small fishery: 135 t in 2009, with a large decline in landings over time. No information is provided on discard patterns, given that discarding in some other plaice stocks is extensive.

Catch-curve estimates of $Z$ varied between 0.6 and 1.2. The estimate for $Z$ appears to be quite variable. These levels of $Z$ are quite high compared to other plaice stocks. There is a possibility that this can be the consequence of declining catchability-at-age due to age-related shifts in distribution and/or seasonal migrations, but is also influenced by an overall decline in catch at all ages over time (see Technical comment 2). The absence of discards numbers-at-age could lead to an underestimation of Z .

In the YPR curve, $\mathrm{F}_{\max }$ is estimated to be 0.24 . Recent values of Z ranged from 0.5 to 1.2 , with $\mathrm{M}=0.12$ this would result in an F of between 0.48 and 1.08 . This is well above $F_{\text {max }}$, however the catch-curve Z's may be biased.

It appears that no survey covers this stock, as no information on this is provided in the Report.

## Technical comments

1 ) Figure 7.11 .6 is described as a catch-curve whereas it is $\log$ catch ratios. Figure 7.11 .8 shows catch-curves.
2 ) A problem with the catch-curve analysis is that the overall catches are declining throughout the period. Hence the numbers-at-age $i+1$, year $y+1$ will decline from the number at year $i$ and age $y$ not only due to mortality but also due to a decline in the overall amount of fishing and catch. This will bias the Zs upwards to some extent. Landings appear more stable from about 2004 onwards, so Z estimates for those years could be shown separately (will probably remain quite high, though).
3 ) Some smoothing of age composition estimates is expected from using an annual ALK built up from all available fleet sampling data. Lumping data to avoid inadequate ALKs at the season, fleet stratum level doesn't necessarily provide better estimates it just covers up the underlying deficiencies and leads to non-quantifiable bias and precision.
4 ) Given a time-series of landings-at-age data, why does the WG resort to catch-curve analysis rather than a simple separable VPA approach with a range of terminal Fs and $S$ values?

5 ) Are there any data on fishing effort of fleets taking the bulk of the plaice catch? Does this suggest any trend in F?

## Conclusions

The RG considers that the analysis could indicate exploitation rates in excess of yield-per-recruit $\mathrm{F}_{\text {msy }}$ proxies; however the magnitude of any difference is unknown without additional information on catchability-at-age and removing any effect of overall reductions in catch.

Improved sampling is needed to allow seasonal ALKs. Estimates of discards-at-age are also needed.

## Plaice in Divisions VIIf,g (Celtic Sea) (Report Section 7.10)

1 ) Assessment type: Update, no change in assessment
2 ) Assessment: analytical
3 ) Forecast: presented
4 ) Assessment model: XSA-tuning two commercial + one survey
5 ) Consistency: The trends and estimates of fishing mortality, SSB and recruitment in this assessment are consistent with last year's assessment

6 ) Stock status: SSB estimate for $2010(1300 \mathrm{t})$ is currently below $\mathrm{B}_{\mathrm{pa}}(1800 \mathrm{t})$ and just above $B_{\lim }(1100 \mathrm{t})$. Recent F estimates are around 0.41 , above the range of $\mathrm{F}_{\text {MSY }}(0.19-0.36)$ proposed by the WG (No Fpa of $\mathrm{F}_{\text {lim }}$ defined).With exception of 1994 year class, all recruitments at age 1 since 1992 have been below the long-term arithmetic mean.
7 ) Management Plan: There is no management plan for Celtic Sea plaice.

## General comments

The WG addressed the ToRs requiring an update assessment and associated management advice, and the assessment was carried out according to the Stock Annex description. The RG found no errors in the implementation of the assessment and forecast, and the results were carried over correctly to the advice sheets.

The WG Report and Stock Annex exclude ecosystem information. The WG Report cites a study on plaice recruitment indicating effects of sea temperature anomalies but does not directly use this information in the assessment or advice (the forecast is considered relatively insensitive to incoming recruitment and hence any environmental effects, although it is acknowledged these could affect longer-term forecasts if they were done).

There is no management plan for this stock.
An important mixed fishery issue is the management of sole and plaice taken in the same fisheries. The sole stock appears healthy whilst the plaice stock is depressed. The WG Reports that increases in mesh size to reduce plaice discarding could (as shown by WGSSDS 2004) also benefit the sole stock. There at present is no attempt to link the advice for sole and plaice taken together in the same fisheries.

As with VIIa plaice, the absence of discards data in the assessment is a serious problem, given the very high discard rates in the 80 mm beam trawl and otter trawl fisheries, and the sexual dimorphism in plaice that may lead to a very high discard rate across a relatively broad age range of males. Information on discard survival is sparse and not referred to by the WG.

The WG estimated total international landings for 2009 were $463 \mathrm{t}, 10 \%$ above the TAC (420 t) and $25 \%$ below the status quo prediction given by last year's assessment $(580 \mathrm{t})$. Discards are considered to be significant but are presently not quantified. In 2009 French data were not available. The retrospective bias in F and abundance could lead to overoptimistic forecasts.

The status quo catch forecast implies that SSB will continue to rise above Blim in 2010, 2011 and in 2012, assuming $\mathrm{GM}_{89-08}$ recruitment levels, but will remain below $\mathrm{B}_{\mathrm{pa}}$.

## Technical comments

1 ) The WG does not explain the derivation of the 2009 landings figures in the absence of official French data.
2 ) The use of the quadratic smoother to estimate weights-at-age should be reviewed in any future benchmark. This year the WG has had to over-write some recent estimates for younger ages due to anomalous results.
3 ) The WG screens survey tuning data for internal consistency and provides Z indices, but does not go as far as deriving recruitment and SSB trends and retrospective analysis as can be provided by SURBA. Given the strong survey catchability trends in XSA (minus discards), the performance of the survey in a stand-alone survey based assessment would be informative.
4 ) The XSA converged after 71 iterations. This indicates some difficulty in finding a solution. Evidence for a power model at ages 1-5 is weak in the beam trawl survey and the decision for this model in the past is probably a patch-up to deal with the lack of discards data.
5 ) As with VIIa plaice, the use of an XSA catch-at-age assessment is an issue when such a large fraction of the catch is missing (discarded), particularly as discarding of males may extend over a relatively wide age range. The full magnitude of the problem has been obscured by the WG not presenting raised discards estimates by age and fleet (ideally including a breakdown by sex) for the series of recent years when such data have been required by the EU Data Collection Framework. The WG must address this as a matter of urgency in preparation for benchmarking of the stock.
6 ) The WG should include in the Stock Annex a description on how the commercial tuning fleet data have been estimated. Is it standardised (vessel power, length, etc)? Does it include all vessels in the fleet or a subset (to be more independent of the catch-matrix)?

## Conclusions

The RG considers that this assessment is not appropriate for providing quantitative management advice, and is useful only for indicating long-tem trends. The assessment has a retrospective bias that probably originates from the absence of the large discard quantities from the assessment. Without any knowledge of how discard rates have changed over time, the ability to implement a combined catch and survey analysis is seriously compromised, and the strong catchability trends induced in the sur-vey-series is indicative of this. The type of retrospective behaviour of the assessment has also been seen in the VIIa plaice assessment; although to a more extreme extent due to the very high discard rates in VIIa. It is important that the ICES review process identifies when update assessments are no longer performing adequately in order that a suitable benchmarking process can be initiated to try and improve the assessment and advice.

The RG proposes that:
1 ) The E\&W BT survey (either raw or modelled) is used for developing an interim fishery-independent assessment of stock trends. It is acknowledged that survey data for the older age classes is relatively sparse, which would need to be allowed for.

2 ) The WG focuses on reconstruction of landings and discards-at-age (by fleet and ideally also sex disaggregated) and that a data compilation and
benchmark process is scheduled at a suitable date to develop a more robust assessment approach. The Benchmark should include VIIa plaice as there are similar problems with the two stocks and there may be mixing between the stocks.

## Plaice in Divisions VIIe (Western Channel) (Report Section 8.2)

1 ) Assessment type: Update, Benchmarked in 2010
2 ) Assessment: Analytical
3 ) Forecast: Short-term forecast provided
4 ) Assessment model: XSA tuned by three commercial fisheries and two surveys. Separable VPA was used for data screening and SURBA for examining tuning series.
5 ) Consistency: The assessment was made according to the settings decided in WKFLAT 2010. Changes to previous assessment were: addition of $15 \%$ of Q1 catches-at-age from VIId; change in first year with catch-at-age data to 1980; reduction in $\mathrm{F}_{\mathrm{bar}}$ from $\mathrm{F}(3-7)$ to $\mathrm{F}(3-6)$; and truncation of FSP survey to exclude age 9 .
6 ) Stock status: Outside safe biological limits. SSB has been below $\mathrm{B}_{\mathrm{pa}}$ since 2004. Estimated fishing mortality declined sharply from around 0.7 in 2006-2008 to 0.44 in 2009, just under $\mathrm{F}_{\mathrm{pa}}(0.45)$ but above the proposed $\mathrm{F}_{\mathrm{msy}}$ of 0.14-0.31.

7 ) Management Plan: None

## General comments

The WG addressed the ToRs relevant to providing advice, and the assessment was carried out according to the WKFLAT Benchmark Assessment. The RG found no errors in the implementation of the assessment and forecast, and the results were carried over correctly to the advice sheets.

The WG Report does not include ecosystem information. The Stock Annex includes reference to impacts of beam trawling. WKFLAT attempted to include sea temperature as an index in the tuning file but this only explained extremes of recruitment and has not been included in the WGCSE assessment.

There is no EU management plan for this stock.
An important mixed fishery issue is the bycatch of plaice in the sole fishery in VIIe, which is subject to the sole management plan. A key finding is that F on both sole and plaice declined sharply by about the same amount between 2008 and 2009 (just under $40 \%$ reduction), although this is greater than the $23 \%$ reduction in beam trawl effort in VIIe between 2008 and 2009. Subsequent assessments may revise the 2009 F upwards.

Discarding occurs at a lower rate than in adjacent plaice stocks. Last year's RG and WG agreed that it would be useful to include discards in the Benchmark. However WKFLAT advised against including noisy discards data as these could degrade the management advice whilst simply rescaling F and SSB series. Discards data are available for 2002-2009 but not included this year. The RG suggests that raised discards estimates should be tabulated and a quantitative evaluation made of the likely fishing mortality rate due to discarding.

Overall, the WG has done a good job with the available data. The migratory effect between the adjacent Area VIId has been partly covered by adding $15 \%$ of the catches from VIId into VIIe. The calculation is based on old tagging data. The RG agrees with the WG that new effort on tagging would be useful for confirming mixing rates. Fur-
thermore, suitably designed tagging experiments could yield valuable data on fishery selectivity.

The changes to the assessment have improved the bias in the retrospective pattern of F but uncertainties still exists in recruitment estimates.

Figure 8.2.9 is missing from the Report.

## Technical comments

1 ) The upward adjustment of SSB throughout the series caused by the addition of VIId data means that the basis for the previous Blim and $\mathrm{B}_{\mathrm{pa}}$ is no longer valid. A revision to the precautionary reference points is needed if these are to be retained for the stock. The Blim value now lies below all historical SSB values in the assessment (see Figure 8.2.11). Alternative reference points are considered in the Stock Annex, but the old values continue to be added to the biomass and stock-recruit plots which is misleading.

## Conclusions

The RG accepts the assessment as a basis for providing quantitative management advice, on the basis of the improved retrospective performance although there still remains uncertainty in recruitment estimates.

The advice based on the MSY framework or MSY transition is acceptable. The main issue is to reduce $F$ to rebuild the SSB.

## Plaice in Division VIla (Irish Sea) (Report Section 6.7)

1 ) Assessment type: Update
2 ) Assessment: Analytical
3 ) Forecast: Short term (not used for advice)
4 ) Assessment model: FLICA tuned with one survey and two biomass indexes
5 ) Consistency: The assessment is biased for SSB, F and recruitment. The updated assessment provides similar estimates to last year's assessment showing a growing SSB and decreasing F since the 1990s.
6 ) Stock status: SSB estimates for 2009 and 2010 are above $B_{p a}(3100 t)$ and F is below $\mathrm{F}_{\mathrm{pa}}(0.45)$
7 ) Management Plan: No.

## General comments

The WG addressed the ToRs in providing an update assessment with associated management advice. The assessment was carried out according to the Stock Annex description with the exception that in the Stock Annex the survey used in tuning UK BT survey has age 1-7, in WG Report Table 6.7.2.4 marked with bold it is age $1-8$. The RG found no errors in the implementation of the assessment and forecast, and the results of the assessment were carried over correctly to the draft advice sheets. No forecast is provided in the advice sheets.

The WG Report and the Stock Annex exclude any ecosystem information.
There is no management plan for the stock.
The WG has not used mixed fishery data in providing advice in support of management.

The Report was very well structured and the necessary information was easy to find. The sections on MSY calculations, uncertainties and bias and recommendations for benchmark showed that the WG has a good idea of what work can be useful for improving the assessment.

Effort has been decreasing in later years and only $1 / 3$ of the TAC is utilised. The fleets targeting plaice or having plaice as an important bycatch have experienced large reductions in effort since the 2000s. Belgian beam effort has declined four-fold since the mid-2000s, and Irish beam trawl effort has halved in the same period. The UK beam trawl fleet has virtually ceased operating in the Irish Sea, and the UK (E\&W) otter trawl fleet which fishes predominantly in parts of the Irish Sea where plaice are most abundant, has also declined to very low effort. As with VIIa sole, international landings of plaice in 2009 were the lowest level in the series (although discard rates are high). A reduction in F on plaice would be expected, and the assessment indicates a progressive reduction over at least the last decade.

Discarding is a very large component of the fishery catches. In 2009, observer data indicated $90 \%$ discarding by number in the UK fleet and $99 \%$ by the Irish fleet. Sexual dimorphism in plaice growth means that a very large proportion of male plaice are probably discarded. Discard survival is poorly known. Given the magnitude of the discard problem, there is clearly no basis for assessing the stock using a catch-at-age model including landings only. There are indications that the selection pattern or/and discard behaviour has changed in the fishery over time.

The outcome of the deficient catch-at-age matrix is that the assessment model cannot reconstruct population numbers that match the trends in survey indices, leading to pronounced retrospective bias and catchability trends in the surveys that have been evident in many of the recent assessments. The catchability trends are most pronounced at ages $2-5$ in the beam trawl survey, covering age classes likely to be most heavily discarded. There are contradictory signals provided by the surveys, which have different spatial coverage.

The ability to implement a catch-at-age model when recent F estimates are around half the M value of 0.12 is debatable, particularly when discard F (if estimated) would be much higher than this. Results of annual egg production method (AEPM) survey estimates for the whole Irish Sea in 1995, 2000, 2006 and 2008 provide SSB estimates 1.5-3.3 times larger than the ICA estimates, and do not suggest a continued steep increase in SSB through the 2000s as indicated by ICA. Discrepancies between recent AEPM and catch-at-age assessment estimates of SSB for North Sea plaice have been resolved by including discards estimates in the assessment.

The egg production survey trend matches the groundfish trawl survey indices of SSB (which cover the whole Irish Sea) more closely as shown in the Figure below. Splitting the egg survey estimates into eastern and western components shows that the increase between 1995 and 2000 is effectively due to the eastern Irish Sea component. The UK (BTS) survey covers only the eastern Irish Sea.


Relative trends in Irish Sea plaice SSB from groundfish surveys and beam trawl survey, with absolute estimates from applications of the annual egg production method ( $\pm 2$ SE) plotted on secondary axis. <Figure included in Advice sheets>

The main indicator of stock status emerging from the catch-at-age assessment is the expansion of the age composition over time, indicating a reduction in $F$ to low values as would be expected given the ratio of fishery landings to egg production estimates of SSB (4-10\% since 2000). However the key issue is the absence of discards estimates, which will also include mature fish. Log catch ratios-at-age for the UK beam trawl survey in the eastern Irish Sea are noisy at the start of the series but show some reduction in Z for most ages (see plots below), but the reduction in Z over time is not as pronounced as may be inferred from the ICA. The mean age in the survey has in-
creased slowly, also indicating a reducing mortality (but could also be influenced by recruitment).

## VIIa plaice UK beam trawl survey data:



It is noted that the TAC is substantially underutilized despite the abundance of the stock. This appears to be due to marketing issues.

## Technical comments

1 ) The landings-at-age data do not provide a suitable basis for an analytical assessment, and therefore the WG assessment update is rejected.
2 ) Given the major uncertainties in the ICA model approach, the WG should provide the results of a survey-only model, to indicate trends in abundance and mortality.
3 ) The RG noted a small discrepancy between stock weight in Table 6.7.2.7 and the stock weight used in the input file (for 2008) although this is largely irrelevant given the rejection of the assessment.

## Conclusions

The RG considers the WG update assessment to be unsuitable as a basis for providing management advice other than as an indicator of the general reduction in F on ages less subject to discarding. The problems with the assessment are analogous to those of the VIIf and g plaice assessment, but are more extreme due to the higher discard rates. The RG recommends that the Advice Sheet should be reworked to provide only trends from surveys and general indicators of stock status such as the broad age range indicating low F (also indicated by the large egg production SSB estimates relative to catches).

On the basis of the overall evidence presented, the RG agrees that a confident statement can be made that the SSB is currently well above the $\mathrm{B}_{\mathrm{pa}}$ value (which is however based on an assessment excluding discards) and that the stock is lightly exploited. Without knowledge of the true level of F at-age due to landings and discards, it is not possible to calculate a meaningful Fmsy value.

The RG recommends that a data compilation and benchmark assessment is scheduled for this stock at a suitable date. The WG has proposed some approaches for a benchmark in the WG Report. It is recommended that the VIIa and VIIf and g plaice stocks
are benchmarked at the same meeting as there are issues common to both stocks. There are also possible stock mixing dynamics to consider. Important tasks for the Benchmark are:

1 ) Compilation of fleet-raised discards estimates for as many years as possible, and an evaluation of the quality of the data based on the ICES QA framework;
2 ) Evaluation of all available survey data including data not presently provided to the WG (e.g. fourth Q western IBTS survey; UK FSP survey data; egg production survey results).

3 ) Better use of survey data to estimate key biological parameters (weight-atage; maturity)
4 ) Disaggregation of the fishery and survey data by sex would be advantageous as the selectivity and discard rates will be different by sex due to slower growth in males. Maturity and natural mortality in males and females may be different.
5 ) Model building using more flexible statistical models to more appropriately account for the different types and magnitude of errors in the data.

## Sole in Division VIIb, c (West of Ireland) (Report Section 7.12)

1) Assessment type: no advice

2 ) Assessment: not presented
3) Forecast: not presented
4) Assessment model:
5) Consistency:
6) Stock status: unkown
7) Management Plan: none.

## General comments

The landings and lpue of Sole in VIIbc appear to have been more or less stable since the start of the logbooks time-series in 1995.

Sole in VIIb are mainly caught by Irish vessels.

## Technical comments

None.

## Conclusions

There are no suitable data on which to base meaningful advice.

## Sole in Divisions VIIh-k (Southwest of Ireland) (Report Section 7.14)

1 ) Assessment type: No assessment, Landings and catch-at-weight sampled for 1993-2009, same advice as last year
2 ) Assessment: None
3 ) Forecast: None
4 ) Assessment model: catch-curve analysis; Yield-per-Recruit analysis
5 ) Consistency: Same approach as last year
6 ) Stock status: The stock is assumed to be within safe biological limits, but the assumption is based on very basic analysis from only parts of the stock unit.
7 ) Management Plan: None.

## General comments

The WG addressed the ToR by updating the catch data and $Z$ estimates.
Area VIIh is considered part of the stock for assessment purposes (management unit) yet it is not believed to be part of the biological stock. There is no possibility to review the background of the stock components since there is no available Stock Annex.

Area VIIh is not sampled although a considerable part of the landings in the assessed stock is from this area. The WG states correctly the inappropriateness of raising landings to all landings in VIIhjk. The conclusion of why VIIh is part of the stock is not defined.

The estimated $Z$ is variable. The $Z$ was peaking in 2003 and on its lowest in 2007. The WG estimate of landings does not show a similar trend but is rather stable in this period.

## Technical comments

1 ) Figure 7.14 .6 is described as a catch-curve whereas it is $\log$ catch ratios.
2 ) As with VIIhk plaice, problem with the catch-curve analysis is that the overall catches are declining throughout the period. Hence the numbers-at-age $1+1$, year $\mathrm{y}+1$ will decline from the number at year I and age y not only due to mortality but also due to a decline in the overall amount of fishing and catch. This will bias the Zs upwards to some extent. Landings appear more stable from about 2004 onwards, so Z estimates for those years could be shown separately (will probably remain quite high, though).
3 ) As with VIIh-k plaice, some smoothing of age composition estimates is expected from using an annual ALK built up from all available fleet sampling data.
4 ) In the yield-per-recruit Section it says $\mathrm{M}=1.0$, should be 0.1
5 ) Given a time-series of landings-at-age data, why does the WG resort to catch-curve analysis rather than a simple separable VPA approach with a range of terminal Fs and $S$ values?
6 ) Are there any data on fishing effort of fleets taking the bulk of the sole catch? Does this suggest any trend in F?

## Conclusions

The RG considers that the analysis could indicate exploitation rates close to yield-perrecruit $\mathrm{F}_{\mathrm{msy}}$ proxies ( $\mathrm{F}_{\mathrm{max}}$ ). However the magnitude of any difference is unknown without additional information on catchability-at-age and removing any effect of overall reductions in catch.

Choice of FMSY proxies should be consistent with other sole stocks.
Improved sampling including seasonal ALKs is needed to make any progress with the quality of the analysis.

## Sole in Division VIIf,g (Celtic Sea) (Report Section 6.7)

1 ) Assessment type: update
2 ) Assessment: analytic
3 ) Forecast: presented
4 ) Assessment model: XSA three tuning fleets two commercial and one survey
5 ) Consistency: Have been some large retrospective adjustments of $F$ and SSB (underestimation of SSB and overestimation of F) but retrospective analysis shows fairly stable results for last few years.
6 ) Stock status: $S S B>B_{\text {trigger }}$ and $F<F_{m s y}$, the 2007 and 2008 year classes are above the long-time average.
7 ) Management Plan: None.

## General comments

The WG Report is well structured and easy to follow. The WG addressed the ToRs relevant to providing an update assessment and associated advice. The assessment was carried out according to the Stock Annex description. The RG found no errors in the implementation of the assessment and forecast, and the results were carried over correctly to the advice sheets.

The WG Report and Stock Annex do not include ecosystem information.
Discarding is estimated to be a minor issue for this stock. However other species including plaice can be discarded in the sole fishery.

An important mixed fishery issue is the management of sole and plaice taken in the same fisheries. The sole stock appears healthy whilst the plaice stock is depressed. The WG Report on Celtic Sea plaice states that increases in mesh size to reduce plaice discarding could (as shown by WGSSDS 2004) also benefit the sole stock. At present there is no attempt to link the advice for sole and plaice taken together in the same fisheries.

The Working Group estimated the total international landings at 790 t in 2009 which is about $20 \%$ below the 2009 TAC ( 993 t ). The TAC is not fully utilized and many age groups are caught in the fishery.

Not all settings are updated in the Stock Annex for last year's assessment.

## Technical comments

1) The divergent signals from the commercial and research beam trawl tun-ing-series are of concern. It is not enough to assume that somehow they cancel each other out (two wrongs don't make a right). Presumably on its own, the survey would produce higher F and lower survivors in recent years. Neither the WG Report nor the Stock Annex describes how the commercial beam trawl tuning data are derived (e.g. is it just annual catch divided by annual effort). It is not possible from the provided data to consider how the Trevose closure, or any changes to fleet structure and activities associated with the large effort reduction in the UK beam trawl fleet, might have affected the lpue of sole in recent years. The WG needs to establish the cause of the difference between the fleets; if it is related to changes in the commercial fishery activity causing more targeting of sole
in coastal waters in recent years the problem may be with the commercial fleet. It is recommended that VMS data is used to evaluate spatio-temporal effects in sole lpue.
2 ) The choice of a Ricker model for the Fmsy evaluation is largely driven by the four largest SSB and associated recruitment, which also happen to be the first four in the time-series. The perception of S-R patterns is therefore affected by time-series effects that are well known to cause bias in fitted stock-recruit curves (see Hilborn and Walters book). The SSB at the start of the series may already be substantially depleted and there is no biological reason to suspect strong density-dependence of recruitment at these SSB values.
3 ) Given the absence of any evidence for reducing recruitment as SSB is reduced, the $\mathrm{F}_{\text {max }}$ and $\mathrm{F}_{\text {msy }}$ values are not surprisingly very close. However, note that the $\mathrm{F}_{\text {msy }} / \mathrm{F}_{\max } \mathrm{F}$ of $\sim 0.3$ leads to a substantial \% depletion of SSB/recruit. An F of $\sim 0.3$ in the 1970s was associated with a declining trend in SSB despite fairly static recruitment. Comparisons of $\mathrm{F}_{\mathrm{MSY}}$ ref points with other sole stocks should be carried out.
4 ) In the advice sheet table "Outlook 2011" the SSB referred to for 2011 is the SSB for 2010 (according to the forecast table in the WG Report). The correct figure is 5050 t .

## Conclusions

The RG agrees that the update assessment is suitable for providing quantitative management advice based on a forecast, as recent assessments appear relatively stable from year to year. However concerns regarding the divergent signals from the two UK beam trawl tuning-series should be investigated further, before any future benchmark.

The stock appears to be in a healthy state and should be exploited in accordance with $\mathrm{F}_{\mathrm{msy}}$. Further work may however be needed before agreeing an Fmsy reference point, including comparison with other sole stocks.

## Sole in Division VIIe (Western Channel) (Report Section 8.3)

1 ) Assessment type: update
2 ) Assessment: analytical
3 ) Forecast: short-term forecast provided
4 ) Assessment model: XSA two commercial tuning fleets, two historic commercial tuning fleets and one survey
5 ) Consistency: Benchmark procedure in 2009 failed to develop an update procedure due to retrospective bias problem.
6 ) Stock status: Biological reference points were rejected by WKFLAT 2009
7 ) Management Plan: A management plan for this stock was agreed in May 2007 (Council Regulation (EC) No 509/2007).

## General comments

The WG addressed the ToRs relevant to providing advice. The WGCSE addressed a special request from the UK to develop an assessment allowing the VIIe sole Management Plan to operate. The RG found no errors in the assessment and forecast, and the results were carried over correctly to the advice sheets apart from some errors in the header of the forecast table which were transmitted to the WG Chairs.

In 2009 WKFLAT benchmarked this assessment, but failed to develop an update procedure, as it was not possible to address the cause of the substantial retrospective bias in F and SSB. The 2010 WGCSE assessment was based on previous accepted XSA formulations described in the Stock Annex, with an decrease in the F-shrinkage SE from 1.0 to 0.5 (stronger shrinkage) and an increase in the time period for shrinkage from 5 to 10 years. A change in $F_{\text {bar }}$ from 3-7 to 3-9 also seems to have been made. The new settings still showed retrospective bias in the past but improved the pattern for recent years.

The WG Report includes brief ecosystem information related to discard rates and the impacts of beam trawling on benthic communities. The WG suggests that sole recruitment patterns may be related to changes in the NAO but does not include any environmental data in the assessment and advice process.

There is an EU management plan for this stock aimed at achieving a long-term F target of 0.27 in three year stepped reductions. WGSSDS $(2005,2006)$ showed that a long-term target F of 0.27 had a low risk of stock depletion below lowest observed values whilst maintaining yield within $10 \%$ of the yield associated with $\mathrm{F}_{\text {max }}$ and average recruitment.

An important mixed fishery issue is the bycatch of plaice in the sole fishery in VIIe. A key finding is that $F$ on both sole and plaice declined sharply by about the same amount between 2008 and 2009 (just under 40\% reduction), although this is greater than the $23 \%$ reduction in beam trawl effort in VIIe between 2008 and 2009. Otter trawl effort has been declining over a longer period. Subsequent assessments may revise the 2009 F upwards on sole and plaice.

Discarding of sole occurs at a very low rate in the fisheries and is not included in the assessment. The stock has a broad age composition, and year classes can be followed in the commercial fishery until at least age 12, indicating the stock has not been too heavily exploited in the past.

A general problem in this year's Working Groups has been the non-availability of official landings data from France, which is generally responsible for about one third of VIIe sole landings. Official international landings in 2009 were 374 t , well below the 2009 TAC ( 650 t). A derivation of French landings from sales slips resulted in the WG estimate of 626 t .

Estimates of $\mathrm{F}_{\mathrm{msy}}$ and its proxies were all considered highly uncertain for this stock and therefore not considered appropriate. The WG has decided to use the results of the stochastic simulations carried out by WGSSDS in 2006 to propose an FMSY of 0.27.

The Report Section needs reviewing to ensure all table and figure numbers are correct and to ensure a consistent naming of tuning fleets in the text and tables.

## Technical comments

1) The two commercial tuning fleets show pronounced negative residuals around 2003-2005, associated mainly with the 1998 year class. The WG notes that the commercial fleets indicated a bigger 1998 year class than was indicated by the survey, and suspects this may be a result of mixing between VIIe and VIIfg occurring beyond the western limit of the survey. The WG should consider spatial mapping of cpue data from the different fleets in the English Channel and VIIfg, linking VMS with logbook data and shore-based and at-sea sampling data.
2 ) The WG states that "recruitment estimates are consistent between the sin-gle-fleet XSA runs, although the final estimates vary slightly". However the recruitment estimates from the survey and commercial fleets diverge substantially in the final years. The XSA diagnostics show that at ages 3-5, the UK beam trawl survey generates much larger survivor estimates and lower F estimates than the commercial fleets, with the differences becoming less pronounced at the older ages. Differences are also apparent in the combined-age cpue where the UK beam trawl survey seems stable but noisy whereas lpue in the beam trawl fishery has been declining continuously since the 1990s. The WG should review the appropriateness of the commercial fleets for providing indices for the younger age classes.
3 ) $\mathrm{F}_{\mathrm{bar}}$ in the Report model settings table and Stock Annex table is age 3-7; however the WG has used an $\mathrm{F}_{\mathrm{bar}}$ of 3-9 in the assessment and advice sheets.
4 ) The short-term predictions previously used $\mathrm{F}_{\text {sq }}$ for the interim year (as area misreporting meant the TAC was not limiting). This year a TAC constraint was used for 2010 on the basis that "recent evidence suggests that the TAC is likely to be observed" although the evidence is not described. In practice, the TAC constraint leads to an F in 2010 of 0.24 which is close to the low F estimate of 0.25 for 2009.
5 ) The SEN file Table 8.3.16 includes a figure for $\mathrm{N}(1)$ of 2815 that does not represent either the XSA estimate or the GM of 4332. This may have no impact on the $\mathrm{F}_{\text {MSY }}$ bootstrap computations.

## Advice sheet

In the Advice sheets Outlook table the SSB for 2011 is given as 2400 t . This is the 2010 SSB. The correct value is 2544 t . The 210 landings should be 618 t not 608 t . ICES was advised of the error.

## Conclusions

The RG agrees that the revised assessment is a pragmatic solution to allow implementation of the Management Plan. However there remain a number of issues with the assessment. The retrospective pattern is now less apparent in the most recent years but is still apparent in the past, and could reappear in future assessments. The RG agrees that a future benchmark should evaluate spatio-temporal dynamics to try and resolve conflicts between data sources. Considerable work would be required prior this to develop the necessary datasets.

The WG has proposed tagging programmes as a means to provide quantitative estimates of stock mixing in plaice in the English Channel. Extending such a programme to include sole in the Channel and VIIfg could also provide valuable data on spatial dynamics of sole as well as information on selectivity patterns in different gears.

## Sole in Division VIIa (Irish Sea) (Report Section 6.8)

1 ) Assessment type: Update
2 ) Assessment: Analytical
3 ) Forecast: Short-term forecast provided
4 ) Assessment model: XSA
5 ) Consistency: 2010 WG assessment is very consistent with the 2009 WG results, continuing to show a reduction in SSB (to lowest in series) and a reduction in $F$ to close to the lowest in the series.

6 ) Stock status: SSB (2009) has fallen to around half of the Blim of 2200 t . F (2009) is just below $\mathrm{F}_{\mathrm{pa}}$ of 0.3.

7 ) Management Plan: There is no management plan in place for Irish Sea sole.

## General comments

The WG addressed the ToRs in providing an update assessment and associated management advice. The assessment was carried out using the procedure detailed in the Stock Annex, with the addition of 2009 catch and survey data. The implementation of the assessment appears to be correct. The advice sheet was not completed at the time of the review.

The WG Report does not include any ecosystem information relevant to VIIa sole and its fisheries.

There is no management plan. ICES recommends that a management plan is put into place given the declining biomass.

Mixed fisheries data are not taken into account by the WG in providing advice in support of management of VIIa sole (Irish Sea Overview Section was not available for review).

Discarding is very low in this stock.
The fleets targeting sole or having sole as an important bycatch have experienced large reductions in effort since the 2000s. Belgian beam effort has declined four-fold since the mid-2000s, and Irish beam trawl effort has halved in the same period. The UK beam trawl fleet has virtually ceased operating in the Irish Sea, and the UK (E\&W) otter trawl fleet which fishes predominantly in parts of the Irish Sea where sole are most abundant, has also declined to very low effort. International landings of sole are at their lowest level in the series. A reduction in F on sole would be expected, and the assessment provides some indication of a reduction in the last few years.

## Technical comments

1) RCT3 input file has wrong data from March BTS compared with Table 6.8.6. This series should be removed from RCT as it has no input to recent year-class forecasts.

2 ) The short-term forecast uses an $\mathrm{F}_{\text {sq }}$ that predicts landings of 439 t in 2010, larger than the TAC of 402 t , and likely to be overoptimistic given the $65 \%$ TAC uptake in 2009 and the dramatic effort reduction in fleets taking sole in recent years.

## Conclusions

The RG considers the updated assessment is suitable for providing management advice in the form of stock trends and short-term forecast. The SSB has been declining due to a run of below average recruitment coupled with $F$ above $F_{p a}$ and Flim for a number of years in the last decade, although the F estimates have dropped below $\mathrm{F}_{\mathrm{pa}}$ in 2009.

## Whiting in Divisions VIIe-k (Report Section 7.15)

1) Assessment type: Update

2 ) Assessment: Indicator of trends only.
3 ) Forecast: Short-term forecast provided but not used in advice sheets.
4 ) Assessment model: FLXSA tuned with two commercial and three survey indices.
5 ) Consistency: The assessment is consistent with last year's update but shows periods of retrospective bias. The bias in SSB has decreased lately.
6 ) Stock status: The SSB estimate for $2010(43 \mathrm{kt})$ is well above $\mathrm{B}_{\mathrm{pa}}(21 \mathrm{kt})$. There is no $\mathrm{Flim}_{\text {lim }}$ or $\mathrm{F}_{\mathrm{pa}}$ defined.
7 ) Management Plan: None

## General comments

The WG addressed the ToR requiring an update assessment and associated management advice. The assessment was carried out according to the Stock Annex description, with the exception of missing French commercial lpue for 2009 in the tuning file which was not updated. The RG found no errors in the implementation of the assessment and forecast. However, both the $\mathrm{F}_{\mathrm{msy}}$ and the short-term predictions are considered very unreliable and are not carried forward to the advice sheets.

The WG Report and the Stock Annex do not contain any information on ecosystem aspects or environmental drivers.

There is no EU management plan for this stock.
The WG has not used mixed fishery data in developing advice for this stock. A brief description is given in the WG Report of fisheries taking whiting. The WG proposal that square mesh panels of 120 mm should be used to reduce haddock discards will impact whiting and all such measures should be viewed in the context of mixed species catches.

The absence of official French landings data is an issue for data quality in 2009, and has also resulted in non-availability of French commercial tuning lpue for 2009. The origin of the WG estimates of French landings in 2009 (associated with large unallocated figure for 2009) is not explained.

As with VIIb-k haddock, the discard issue for the whiting stock is considerable and is discussed at length by the WG, which excluded discards data from the assessment. Discards data of variable coverage are provided to the WG but are not used for the following reasons given by the WG: 1) don't have discards data for the full period of landings-at-age data; 2 ) sampled fleets are not representative of the main fleets in the fishery and 3) need to examine and agree the best raising procedures for the various fleets. In the "recommendation for the next Benchmark" section, the WG at least makes a clear proposal for work that needs to be done to make use of the discards data. The same proposal should apply to Area VII haddock and other stocks with significant discarding not included in the assessment. The WG should ensure that these proposals are followed up and a commitment is made to resolving the discards data issues.

The TAC has not been utilized in many years and in 2009 only $1 / 3$ of the TAC was landed. However the stock area does not fully correspond to the assessment area.

## Technical comments

1 ) The internal consistency of the surveys is generally quite poor. Some of the year-class time-series in the SURBA data screening plots in Figure 7.15.9 look out of synch by a year, e.g. age 3 in the bottom plot. Is this indicative of age errors?
2 ) The tuning fleets IRGFS7 used in the assessment shows a strong increasing trend in time; it has very large residuals and not very good internal consistency for age 3 and older ages. The concern is that it has a very high weighting in the assessment for most ages.
3 ) For the short-term forecast, the WG has accepted the 2009 year-class estimate from XSA (despite poor precision) but has reduced the 2008 estimate by a rather arbitrary $25 \%$ based on historic adjustments to the 1999 year class. However note that the current estimate for the 1999 year class is close to the current XSA estimate for the 2008 year class. If this forecast was actually being used to set a TAC, the $25 \%$ adjustment would feed directly into the 2011 TAC and would be hard to defend to the industry.
4 ) The estimation errors in the catch data make XSA a questionable model for this assessment. As with VIIb-k haddock, the RG suggests exploring a more statistical model that can deal with a variety of different datasets of differing quality and more accurately deal with the types and magnitude of errors.
5 ) As with VIIb-k haddock, the WG has declined to put forward any candidate $\mathrm{F}_{\text {MSY }}$ values. The reasons given in this case are the uncertainty in the data and the absence of discards mortality estimates. Fmax is clearly unsuitable as the YPR is asymptotic. The YPR reference points $\mathrm{F}_{0.1}$ and $\mathrm{F}_{35 \%} \%-40 \%$ appear well estimated and are in the range $0.16-0.24$ but the analysis excludes discard mortality. The WG is in the best position to propose appropriate long-term F targets for this stock and should continue to explore possibilities through simulation.

## Conclusions

The RG supports the use of the updated XSA assessment as providing an indication of longer-term trends but not for providing a forecast. The assessment shows difficulties in estimating recruitment. The estimate of the strong year class 2008 is shifted down by $55 \%$ in this year's assessment. The downshift for the 2007 year class in last year's assessment was $49 \%$. The indication is nevertheless that the last three year classes are strong, but not of the magnitude as previously estimated.

The stock is currently well above the $\mathrm{B}_{\mathrm{pa}}$ of 21000 t
The major problems to be addressed are the treatment of discards and the difficulties in estimating recent year-class strength based on the surveys.

## Whiting in Division VIIa (Irish Sea) (Report Section 6.6)

1) Assessment type: Update

2 ) Assessment: survey trends
3 ) Forecast: not presented
4 ) Assessment model: SURBA
5 ) Consistency: Retrospective SURBA runs show consistent trends compared to previous assessments.
6 ) Stock status: The state of the stock is unknown; however the stock is perceived to be subject to high fishing mortality and is at an extremely low level. Existing biological reference points are from XSA assessments no longer considered valid.
7 ) Management Plan: No management plan has been agreed or proposed.

## General comments

The WG has addressed the ToR in providing an updated assessment using a survey based model, able to provide information on relative trends. The SURBA model has been implemented using the approach outlined in previous assessments. The Stock Annex does not tabulate SURBA model settings.

The whiting fishery for human consumption in VIIa has effectively disappeared and most of the catch is now discarded in the small mesh fisheries. A range of discards data is presented by the WG but the data are patchy and there is no unified set of dis-cards-at-age for the full period up to 2009. Some fleets do not appear to have discards data after 2002 despite the DCF requirement to collect discards data for stocks with significant discard rates. It is therefore not possible to evaluate the full extent of dis-carding-at-age. It is important that steps are taken to obtain robust and reliable estimates of discards; otherwise the effectiveness of technical measures to reduce discarding will be difficult to quantify.

The TAC in 2010 was set at a very small value of 157 t , but even this was not fully utilised.

The Irish Sea whiting stock is primarily caught by otter trawlers which utilize two main mesh size ranges, $70-89 \mathrm{~mm}$ and $100-119 \mathrm{~mm}$. Effort of trawlers utilizing the larger mesh range, traditionally targeting whitefish (cod, haddock, whiting) has seen a large declined since 2003. The smaller range however has remained relatively stable. There primary target species is Nephrops from which whiting is discarded at a high rate.

In late 2009, a number of Irish vessels operating within the Irish Sea Nephrops fishery incorporated a Swedish grid into otter trawls, as part of the cod long-term management plan. It is expected that this will reduce the whiting catches of these vessels by $60 \%$ in weight. Furthermore, a small number of vessels began utilizing an inclined separator panel expected to reduce whiting catch by $76 \%$ in weight.

## Technical comments

1 ) Section 6.6.2: Survey data. March groundfish surveys are available from 1992 to March 2010; however in Figure 6.6.2 data are only plotted to 2009.
2 ) The Stock Annex is several years out of date and should be updated before next year's WG.

3 ) The decline in catch rates >MLS appears particularly sharp between 2003 and 2004 in the eastern Irish Sea. The tuning file for the March-East survey suggests this occurs over the age range, indicating a shift in whiting distribution or a change in catchability. The WG should evaluate causes for this.
4 ) The NIGFS East-March tuning data has 1993-2010 in the year range, but the column of years starts at 1994.

## Conclusions

The RG considers that the SURBA model indices of abundance provide credible information showing a severe decline in biomass of whiting in the Irish Sea since the mid-1990s. Available discards data show that most of the whiting catch is discarded but there is no single coherent set of fleet-raised discards data for tracking changes in discarding over time.

## Whiting in Division VIb (Rockall)

Report Section not available.

## Whiting in Division Vla (West of Scotland) (Report Section 3.4)

1 ) Assessment type: Update
2 ) Assessment: Assessment of trends only.
3 ) Forecast: None
4 ) Assessment model: SURBA, with comparative assessment using XSA
5 ) Consistency: The stock has not been assessed since 2007.
6 ) Stock status: It is not possible to evaluate stock status relative to reference points. SSB and recruitment appear to have declined to a very low level. Survey and fishery data indicate declining mortality since the mid-2000s.
7 ) Management Plan: None.

## General comments

The WG met the ToRs by providing an update survey based assessment. The assessments were conducted following the procedures outlined in the Stock Annex, and appear to have been implemented correctly. The correct figures have been carried into the Advice sheets.

There are no ecosystems considerations described although reference is made to possible shifts in predation or regime shifts that could explain discrepancies between SURBA and XSA trends during the 1980s and 1990s.

Mixed fisheries is described as a problem as whiting is mainly linked with fisheries for cod and haddock in VIa that are affected by the cod management plan. Shifts in effort to small mesh Nephrops fisheries would worsen the exploitation pattern for whiting.

There is no management plan for VIa whiting.
Experimental runs using both in SURBA and XSA were performed to evaluated different possibilities for the assessment. The surveys available all have some quality issues, in many cases noisy, or changed and discontinued. Only the ScoGFSQ1 is used for the final SURBA run. However, the trends given by the Q1 and Q4 survey are qualitatively similar and lead to a similar conclusion regarding recent stock trends. The VIa whiting assessment suffers a similar problem to the North Sea whiting assessment in that survey data and XSA results agree for a recent period but diverge considerably in earlier years. The North Sea WG considers this could be due to bias in catch estimates, changes in survey catchability, or changes in natural mortality due to predation or regime shift. It is therefore not possible to consider biological reference points based on long-term XSA results.

Substantial discarding occurs in this stock. This may reduce substantially due to the mesh increase to 120 mm for whitefish vessels. The WG states that Scottish discards are being reworked. It is noted by the RG that the cited methodology for this nonetheless dates back to 2004. It is important that discards data are fully worked up by country and fleet prior to any benchmark assessment.

## Technical comments

1 ) The WG states that weights-at-age from the commercial catch data were used to provide stock weights-at-age for this year's assessment. It is not clarified why the surveys don't deliver any reliable stock weights.

## Conclusions

The RG considers that the updated survey analysis are suitable for providing advice on recent stock trends since the 1990s, but there are some difficulties in interpreting stock trends for earlier years. Even given the uncertainties in the performance of the assessment the stock is likely to be on its lowest level.

The WG outlines several feasible ways to improve the data and assessment before benchmarking and making a formal analytical assessment. The RG recommends that benchmarking of North Sea and west of Scotland whiting should take place at the same time as there appear to be similar problems with the long-term data for both stocks, and linkages between whiting in the two areas could be evaluated.

## Whiting in Division VIIb, c (West of Ireland) (Southwest of Ireland)) (Report Section 7.16)

1 ) Assessment type: No advice
2 ) Assessment: not presented
3 ) Forecast: not presented
4 ) Assessment model:
5 ) Consistency:
6 ) Stock status: unknown
7 ) Manageement Plan: none.

## General comments

Landings are very small 78 t and only a landing table is provided. It appears that France normally has landings for this stock, but no official data are available for 2009. This has however, not been mentioned.

Landings have been decreasing in time, from more than 2000 t in the mid-1990s to below 100 t at present. There are no indications from the WG whether this could be stock or market related. Some of the decline is due to the decreasing landings from France, but a large stock decline would be consistent with the picture for some other whiting stocks in the NE Atlantic.

## Technical comments

None.

## Conclusions

There is no scientific basis for providing advice for this stock.

## Stock Annex 3.2: Cod in VIa

- Stock Annex 3.2 Cod VIa: for latest update see WGCSE 2009, Annex 03.2 Cod VIa


## Stock Annex 3.3: Haddock in VIa

- Stock Annex 3.3 Haddock in VIa: for latest update see WGCSE 2009 Annex 03.3 Haddock VIa


## Stock Annex 3.4: Whiting in Area VI

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Whiting (Area VI) |
| :--- | :--- |
| Working Group | Assessment of Northern Shelf Demersal Stocks |
| Date | 17 May 2007 |
| Last updated | 25 May 2010 (a.jaworski@marlab.ac.uk) |

## A. General

## A.1. Stock definition

Whiting occur throughout northeast Atlantic waters, in a wide range of depths, from shallow inshore waters down to 200 m . Adult whiting are widespread throughout Area VIa, while high numbers of juvenile fish occur in inshore areas. Whiting are less common in Division VIb, and it is likely these fish are migrants from VIa, rather than a separate stock.

While an exploration of stock identity in the North Sea has been carried out, stock definition in Area VI and surrounding waters remains poorly defined (ICESSGISIMUW, 2005). Tagging experiments on recruiting fish have shown that whiting stocks west of Ireland are distinct from those in the Minches, Clyde and the Irish Sea. On the basis of preliminary results from FRS project MF0464, there appears to be three putative populations of whiting are found in VIa, between which interchange is limited. These are along the northwest of Scotland, the Stanton Bank region and the Firth of Clyde. Maximum likelihood analysis indicates a high degree of mixing for adult whiting between IVa whiting and the VIa component off the northwest of Scotland. Within VIa, there was little indication of interaction between population components in the south and that off the northwest coast.

## A.2. The fishery

The demersal fisheries in Division VIa are predominantly conducted by otter trawlers fishing for cod, haddock, anglerfish and Nephrops, with bycatches of whiting, saithe, megrim, lemon sole, ling and a number of skate species. Since 1976, effort by Scottish heavy trawlers and seiners has decreased. Light trawler effort has declined rapidly since 1997 after a long-term increasing trend. More recently, days-at-sea limitations associated with the cod recovery plan and the seasonal closure of some areas has lead to some switching of effort away from VIa.

The demersal whitefish fishery in Area VI occurs largely in Division VIa with the UK, Ireland, Spain and France being the most important exploiters. Landings from Rockall (Division VIb) are generally less than 10 t . The whiting fishery in VIa is dominated by the UK (Scotland) and Irish fleets. French whiting landings have declined considerably since the late 1980s.

Landings of whiting in Division VIa are affected by emergency measures introduced in 2001 as part of the cod recovery programme. Council Regulation $423 \backslash 2004$ introduced a cod recovery plan affecting division VIa. The measures only take effect, however east of a line defined in Council Regulation No $51 \backslash 2006$. Measures brought in in 2002, such as a switch from 100 to 120 mm mesh codends at the start of 2002 (Commission Regulation EC2056/2001), are likely to have had some impact on whit-
ing.The UK implemented a regulation requiring the fitting of a square mesh panel in certain towed gears.
Most catch of whiting comes in non-whiting directed fisheries, particularly the Nephrops trawl fishery. The Nephrops trawl fishery in VIa discards significant amounts of small whiting, making whiting landings figures a poor indicator of removals due to fishing. The proportion of whiting discarded has been very high and appears to have increased in recent years. Whiting also has a low market demand, which contributes to increased discarding and high-grading.
The minimum landing size of whiting in the human consumption fishery in this area is 27 cm .

There has been some problems regarding area misreporting of Scottish landings during the early 1990s, which are linked to area misreporting of other species such as haddock and anglerfish into Division VIb. More recently there has been area misreporting of anglerfish from VIa to IVa, which may have affected the reliability of whitings landings distribution.

## A.3. Ecosystem aspects

No information.

## B. Data

## B.1. Commercial catch

Monthly length frequency distribution data were available from Scotland for Area VIa. A total international catch-at-age distribution for Division VIa was obtained using the raising procedure described in Section 2.3 to raise this distribution to the WG estimates of total international catch from this area. Landings officially reported to ICES were used for countries not supplying estimates directly to the WG. The Scottish market sampling length-weight relationships (given below) have been used to raise the sampled catch-at-length distribution data Working Group estimates of total landings for Division Via.

| Month | b | a |
| :---: | :---: | :---: |
| 1 | 2.9456 | 0.01 |
| 2 | 2.9456 | 0.0094 |
| 3 | 2.9456 | 0.009 |
| 4 | 2.9456 | 0.0088 |
| 5 | 2.9456 | 0.0088 |
| 6 | 2.9456 | 0.0089 |
| 7 | 2.9456 | 0.009 |
| 8 | 2.9456 | 0.0092 |
| 9 | 2.9456 | 0.0095 |
| 10 | 2.9456 | 0.0096 |
| 11 | 2.9456 | 0.0097 |
| 12 | 2.9456 | 0.0097 |

Discard age-compositions are generally available from both Scotland and Ireland, but in recent years (2006 and 2007) lack of access to fishing vessels by Irish observers has meant that no Irish data have been collected. Work is underway to revise the Scottish discard estimates with an aim to reduce bias and increase precision. Such revisions are particularly important for the estimation of total catch for this stock which has
very high discards across a wide age range. A working document set out the methodology of this work at the 2004 meeting of WGNSDS (Fryer and Millar, 2004).

## B.2. Biological

Natural mortality is assumed to be constant ( $\mathrm{M}=0.2$, applied annually) for the whole range of ages and years.

A combined sex maturity is assumed, knife-edged at age 2. The use of a knife-edged maturity ogive has been a source of criticism in previous assessments. However, recent research on gadoid maturity conducted by the UK (NI) gives no evidence for substantial change in whiting maturity since the 1950s, although there has been an increase in the incidence of precocious maturity-at-age 1, particularly in males, since 1998, in the Irish Sea.

As in previous years, SSB is computed at the start of each year, and the proportions of $M$ and $F$ before spawning were set to zero. Stock weights are calculated using a procedure first described in the 1998 Working Group report. To derive representative stock weights for the start of the year for year $i$ and age $j$ the following formula is adopted:
$(C W i, j+C W i+1, j+1) / 2=S W$ at start of year.

## B.3. Surveys

Four research vessel survey-series for whiting in VIa were available to the Working Group in 2007. In all surveys listed the highest age represents a true age not a plus group.

- Scottish first-quarter west coast groundfish survey (ScoGFSQ1): ages 1-7, years 1985-2010.

The survey gear is a GOV trawl, and the design is a minimum of one station per rectangle, but with more depending on logistic limitations. Ages are reported from 0 to the maximum obtained. Sex/Maturity-Sex and Maturity (ICES 4-stage scale) are reported. The Scottish groundfish survey has been conducted with a new vessel and gear since 1999. The catch rates for the series as presented are corrected for the change on the basis of comparative trawl haul data (Zuur et al., 2001).

- Irish fourth-quarter west coast groundfish survey (IreGFS): ages 0-5, years 1993-2002.

The Irish quarter four survey was a comparatively short series, was discontinued in 2003 and has been replaced by the IRGFS.

- Scottish fourth quarter west coast groundfish survey (ScoGFSQ4): ages 08, years 1996-2009.

The Scottish quarter four survey was presented to the WG for the first time in 2007.

- Irish fourth quarter west coast groundfish survey (IRGFS); ages 0-6, years 2003-2009.

This survey used the RV Celtic Explorer and is part of the IBTS coordinated western waters surveys. The vessel uses a GOV trawl, and the design is a depth stratified survey with randomised stations. Effort is recorded in terms of minutes towed. There were 41 stations sampled in 2003, 44 in 2004 and 34 in 2005, corresponding to 1229, 1321 and 1010 minutes towed.

Further descriptions of these surveys and distribution plots of whiting catch rates obtained on these surveys can be found in the IBTS WG Report of 2008.

The indices are provided in Table B.1.
The distribution of catches per unit of effort from the surveys in 2008 are given in Figure B. 1 for the Scottish fourth quarter west coast groundfish survey (ScoGFSQ4); and Figure B. 2 for the first quarter west coast groundfish survey (ScoGFSQ1).

## B.4. Commercial cpue

Due to a number of concerns regarding the non-mandatory recording of effort in terms of hours fished, the present assessment of the stocks does not make use of commercial catch per unit of effort data. The data are included here for completeness (Table B.2) and include:

- Scottish light trawlers (ScoLTR): ages 1-7 years 1965-2005
- Scottish seiners (ScoSEI): ages 1-6 years 1965-2005
- Scottish Neprhops trawlers (ScoNTR): ages 1-6 years 1965-2005
- Irish Otter Trawlers (IreOTB): ages 1-7 years 1995-2005

Data to update these time-series were not available for 2006 or 2007.

## B.5. Fecundity

Fecundity data for a number of areas are available from Hislop and Hall (1974), and was estimated at 4.933 L $^{3.25}$ for whiting in Area VI.

## C. Historical stock development

Whiting has never been a particularly valuable species and has tended not to be targeted by commercial fishermen. It tends to be taken more as a bycatch, with other species fished more intensively in Division VIa, such as haddock, cod and angler fish. As with other gadoids in VIa, whiting stocks have declined steadily since the late 1970s.

## D. Short-term projection

Not done.

## E. Medium-term projections

No medium-term projections are carried out for this stock.
F. Yield and biomass-per-recruit/long-term projections

Not done.

## G. Biological reference points

Precautionary approach reference points:
VIa-"Long-term information on the historical yield and catch composition all indicate that the present stock size is low. A survey-based assessment covering the more recent period indicates that the stock is at its lowest level over this time period. Total mortality is at the highest level over the time period.

ICES considers that $\mathrm{Blim}_{\text {lim }} 16000 \mathrm{t}$ and $\mathrm{B}_{\mathrm{pa}}$ be set at 22000 t . ICES proposes that $F_{l i m}$ is 1.0 and $F_{p a}$ be set at 0.6 ."

VIb-"Landings of whiting from Division VIb are negligible. No assessment has been carried out on this stock."

## H. Other issues

None.

## I. References

J. R. G. Hislop. 1975. The breeding and growth of whiting, Merlangius merlangus, in captivity. J. Cons. int. Explor. Mer, 36(2): 119-127.

Hislop, J. and Hall, W. 1974. The fecundity of whiting, Merlangius merlangius (L.), in the North Sea, Minch and at Iceland. J. Cons. int. Explor. Mer, 36(1): 42-49.

ICES. 2000. ICES CM 2000/ACFM:1.
ICES-SGSIMUW. 2005. Report of the Study Group on Stock Identity and Management Units of Whiting. ICES CM 2005/G:03.

Table B.1. Available survey tuning-series. For IreGFS, effort is given as minutes towed, numbers are in units.

| SCOGFSQ1: Scottish Groundfish Sruvey - Effort in hours - Numbers at age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
| Year | (hours) | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| 1985 | 10 | 3140 | 1792 | 380 | 85 | 23 | 156 | 18 |
| 1986 | 10 | 1456 | 1526 | 403 | 68 | 10 | 9 | 10 |
| 1987 | 10 | 6938 | 1054 | 584 | 143 | 36 | 2 | 1 |
| 1988 | 10 | 567 | 3469 | 653 | 189 | 42 | 5 | 1 |
| 1989 | 10 | 910 | 505 | 586 | 237 | 48 | 3 | 0 |
| 1990 | 10 | 1818 | 572 | 122 | 216 | 61 | 4 | 1 |
| 1991 | 10 | 3203 | 277 | 298 | 22 | 39 | 9 | 1 |
| 1992 | 10 | 4777 | 1597 | 410 | 517 | 56 | 18 | 0 |
| 1993 | 10 | 5532 | 6829 | 644 | 91 | 30 | 11 | 2 |
| 1994 | 10 | 6614 | 2443 | 1487 | 174 | 56 | 15 | 6 |
| 1995 | 10 | 5598 | 2831 | 1160 | 370 | 70 | 17 | 32 |
| 1996 | 10 | 9384 | 2238 | 635 | 341 | 135 | 30 | 5 |
| 1997 | 10 | 5663 | 2444 | 1531 | 355 | 102 | 17 | 4 |
| 1998 | 10 | 9851 | 1352 | 294 | 195 | 50 | 14 | 1 |
| 1999 | 10 | 6125 | 4952 | 489 | 103 | 16 | 1 | 0.4 |
| 2000 | 10 | 12862 | 471 | 152 | 34 | 10 | 11 | 0 |
| 2001 | 10 | 4653 | 1954 | 242 | 41 | 8 | 1 | 1 |
| 2002 | 10 | 5542 | 1028 | 964 | 86 | 15 | 1 | 1 |
| 2003 | 10 | 6934 | 746 | 436 | 300 | 32 | 2 | 4 |
| 2004 | 10 | 5888 | 1566 | 189 | 131 | 44 | 9 | 1 |
| 2005 | 10 | 1308 | 723 | 183 | 35 | 8 | 11 | 2 |
| 2006 | 10 | 1441 | 466 | 282 | 77 | 0.3 | 3 | 0.6 |
| 2007 | 10 | 614 | 522 | 127 | 75 | 16 | 3 | 2 |
| 2008 | 10 | 593 | 127 | 77 | 26 | 8 | 3 | 0 |
| 2009 | 10 | 906 | 387 | 103 | 105 | 20 | 9 | 7 |
| 2010 | 10 | 3523 | 340 | 108 | 52 | 40 | 4 | 3 |


| IR-WCGFS : Irish West Coast GFS (VIa) - Effort (min. towed) - Whiting number at age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
| Year | (min) | 0 | 1 | 2 | 3 | 4 | 5 |  |
| 1993 | 2130 | 14403 | 32643 | 11419 | 1464 | 231 | 13 |  |
| 1994 | 1865 | 264 | 11969 | 4817 | 2812 | 78 | 57 |  |
| 1995 | 2026 | 34584 | 5609 | 6406 | 734 | 186 | 80 |  |
| 1996 | 2008 | 376 | 7457 | 3551 | 374 | 232 | 5 |  |
| 1997 | 1879 | 1550 | 13865 | 8207 | 1022 | 524 | 50 |  |
| 1998 | 1936 | 1829 | 4077 | 3361 | 663 | 121 | 5 |  |
| 1999 | 1914 | 3337 | 3059 | 1965 | 322 | 11 | 12 |  |
| 2000 | 1878 | 682 | 10102 | 2126 | 109 | 109 | 4 |  |
| 2001 | 965 | 1118 | 5201 | 2903 | 149 | 70 | 3 |  |
| 2002 | 796 | 594 | 8247 | 9348 | 820 | 280 | 0 |  |

(cont). Whiting in VIa. Available survey tuning-series. For ScoGFSQ4, numbers are standardised to catch-rate per 10 hours. " + " indicates value less than 0.5 after standardising.

|  | IRGFS: Irish groundfish survey - effort in minutes - numbers at age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort | Age |  |  |  |  |  |  |
| Year | (min) | 0 | 1 | 2 | 3 | 4 | 5 | 6 |
| 2003 | 1127 | 1101 | 12886 | 2894 | 512 | 290 | 102 | 1 |
| 2004 | 1200 | 6924 | 3114 | 1312 | 104 | 35 | 16 | 1 |
| 2005 | 960 | 910 | 2228 | 1126 | 91 | 5 | 4 | 0 |
| 2006 | 1510 | 99 | 1055 | 921 | 214 | 27 | 3 | 0 |
| 2007 | 1173 | 138 | 1989 | 2380 | 722 | 169 | 251 | 122 |
| 2008 | 1135 | 24 | 4342 | 1328 | 573 | 243 | 123 | 36 |
| 2009 | 1378 | 16906 | 1430 | 989 | 325 | 68 | 21 | 41 |

ScoGFSQ4 : Quarter four Scottish groundfish survey - Effort in hours - numbers at age

| Effort |  |  |  |  |  |  |  |  | Age |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | (hours) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1996 | 10 | 5154 | 1908 | 1116 | 570 | 188 | 51 | 6 | 1 | 0 |
| 1997 | 10 | 8001 | 2869 | 951 | 323 | 160 | 46 | 12 | 1 | 0 |
| 1998 | 10 | 1852 | 2713 | 1124 | 149 | 100 | 20 | 1 | 0 | + |
| 1999 | 10 | 8203 | 2338 | 582 | 141 | 33 | 24 | 1 | 1 | 0 |
| 2000 | 10 | 4434 | 4055 | 789 | 160 | 9 | 7 | 1 | 0 | 0 |
| 2001 | 10 | 9615 | 1957 | 1420 | 155 | 40 | 12 | 2 | 0 | 0 |
| 2002 | 10 | 14658 | 1591 | 621 | 479 | 30 | 9 | 5 | 0 | 0 |
| 2003 | 10 | 9932 | 3446 | 567 | 338 | 83 | 27 | 4 | 0 | 0 |
| 2004 | 10 | 5923 | 1758 | 940 | 83 | 57 | 62 | 1 | 0 | 0 |
| 2005 | 10 | 2297 | 308 | 318 | 76 | 9 | 4 | 0.9 | 0.7 | 0 |
| 2006 | 10 | 415 | 296 | 140 | 101 | 35 | 8 | 3 | 0.5 | 0 |
| 2007 | 10 | 1894 | 434 | 326 | 99 | 83 | 48 | 0.6 | 0 | 0 |
| 2008 | 10 | 2297 | 208 | 78 | 110 | 28 | 24 | 4 | 0 | + |
| 2009 | 10 | 4833 | 236 | 178 | 50 | 58 | 12 | 6 | 6 | 0 |

Table B.2. Commercial cpue tuning-series available to whiting in VIa.

| 2009 WHITING AREA 6A |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 108 |  |  |  |  |  |  |  |
| SCOLTR: Scottish Light Trawl: Effort in hours: Numbers-at-age (thousands) |  |  |  |  |  |  |  |
| 1965 | 2005 |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |
| 37387 | 2011.623 | 469.253 | 3512.923 | 393.473 | 14.925 | 5.445 | 0.909 |
| 40538 | 1036.117 | 926.485 | 162.985 | 5508.27 | 333.46 | 32.68 | 6.196 |
| 80916 | 2539.797 | 4967.604 | 1637.023 | 101.256 | 2456.915 | 133.979 | 12.466 |
| 65348 | 1931.014 | 3404.448 | 1868.458 | 677.298 | 51.295 | 844.125 | 58.939 |
| 106856 | 46.897 | 8823.442 | 2211.584 | 578.006 | 278.879 | 28.188 | 516.892 |
| 129741 | 94.958 | 5275.823 | 8514.611 | 712.848 | 143.241 | 35.554 | 3.428 |
| 137728 | 1566.57 | 4472.064 | 1026.561 | 9818.08 | 337.772 | 63.477 | 25.237 |
| 154288 | 13450.885 | 4637.042 | 1716.159 | 334.786 | 5435.152 | 309.86 | 29.756 |
| 93992 | 4613.649 | 12778.492 | 680.372 | 148.997 | 42.975 | 478.522 | 39.083 |
| 88651 | 7452.711 | 15917.02 | 1773.837 | 159.241 | 17.112 | 6.477 | 78.812 |
| 132353 | 10597.964 | 6684.991 | 10431.537 | 837.283 | 79.71 | 12.155 | 2.811 |
| 139225 | 10858.324 | 15481.895 | 3550.826 | 5483.438 | 412.525 | 13.045 | 4.668 |
| 143574 | 18222.115 | 4276.619 | 5983.177 | 773.244 | 1126.782 | 74.579 | 1.916 |
| 127387 | 9805.191 | 5887.935 | 1561.61 | 1814.903 | 127.832 | 244.126 | 3.76 |
| 99803 | 1846.163 | 9530.148 | 2446.896 | 368.018 | 290.896 | 31.887 | 57.01 |
| 121211 | 1856.938 | 4385.272 | 4359.469 | 1052.873 | 170.989 | 172.29 | 10.997 |
| 165002 | 983.137 | 13544.1 | 4617.56 | 1330.75 | 504.711 | 152.752 | 62.619 |
| 135280 | 8248.806 | 2593.129 | 10934.792 | 1899.759 | 316.934 | 74.891 | 62.409 |
| 112332 | 4809.036 | 4322.894 | 2548.597 | 8292.216 | 1696.241 | 253.9 | 54.475 |
| 132217 | 29865.064 | 4084.418 | 2582.188 | 1149.781 | 5206.862 | 592.972 | 221.473 |
| 142815 | 9243.535 | 11577.551 | 2515.313 | 663.96 | 360.662 | 917.939 | 82.73 |
| 126533 | 3187.288 | 6006.487 | 2693.592 | 621.738 | 98.497 | 50.635 | 93.945 |
| 131720 | 12328.429 | 6004.925 | 2767.12 | 1229.144 | 147.776 | 43.178 | 32.132 |
| 158191 | 5358.52 | 15325.219 | 2988.119 | 1334.433 | 316.668 | 46.956 | 2.997 |
| 217443 | 3161.234 | 1640.767 | 5226.339 | 1473.139 | 434.728 | 129.89 | 14.252 |
| 169667 | 4110.42 | 4152.38 | 972.043 | 1380.502 | 386.872 | 51.478 | 6.092 |
| 209901 | 7018.52 | 2968.053 | 3981.784 | 336.752 | 423.153 | 73.429 | 5.829 |
| 189288 | 9761.596 | 6548.587 | 1727.049 | 2100.437 | 113.974 | 102.439 | 10.66 |
| 189925 | 2623.886 | 10105.623 | 4392.988 | 1169.932 | 1701.769 | 51.678 | 46.841 |
| 174879 | 3251.43 | 6503.608 | 5363.793 | 1739.967 | 333.927 | 291.821 | 13.881 |
| 175631 | 1775.509 | 5661.947 | 5310.813 | 1995.375 | 569.453 | 114.177 | 107.935 |
| 214159 | 2738.034 | 8043.865 | 4647.63 | 2543.265 | 833.461 | 213.15 | 24.196 |
| 179605 | 3107.284 | 3973.701 | 5098.515 | 1858.52 | 532.696 | 95.153 | 39.379 |
| 142457 | 3997.939 | 3171.019 | 2547.76 | 2327.54 | 654.589 | 149.808 | 79.812 |
| 98993 | 559.916 | 3273.961 | 1709.217 | 814.593 | 793.265 | 122.037 | 34.883 |
| 76157 | 4363.101 | 2324.771 | 2202.561 | 627.094 | 169.833 | 201.883 | 8.678 |
| 35698 | 575.281 | 2603.626 | 1358.595 | 783.414 | 117.804 | 37.996 | 5.442 |
| 15174 | 389.652 | 848.153 | 1566.132 | 374.617 | 166.509 | 16.845 | 5.038 |
| 9357 | 565.293 | 207.507 | 273.115 | 578.307 | 100.052 | 41.916 | 0.206 |
| 7116 | 1769.901 | 1215.938 | 242.922 | 199.9 | 221.001 | 27.997 | 3.138 |
| 3063 | 217.522 | 400.094 | 268.966 | 23.085 | 27.158 | 14.318 | 2.462 |

Table B.2. continued.

| SCOSEI: Scottish Seine: Effort in hours: Numbers-at-age (thousands) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 2005 |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |
| 1 | 6 |  |  |  |  |  |
| 153103 | 8570.938 | 4534.63 | 19453.707 | 1412.984 | 62.399 | 15.334 |
| 156511 | 2872.249 | 12671.39 | 1491.149 | 13027.566 | 736.15 | 68.22 |
| 158208 | 7058.77 | 23604.969 | 5804.573 | 363.182 | 5528.921 | 304.951 |
| 150094 | 11817.932 | 14128.65 | 4897.227 | 1409.535 | 134.705 | 1651.222 |
| 140718 | 1314.237 | 19167.426 | 4024.433 | 1038.908 | 420.643 | 45.006 |
| 95629 | 979.255 | 2065.056 | 9177.95 | 815.703 | 176.987 | 51.144 |
| 98748 | 3280.938 | 6459.36 | 2466.983 | 14808.06 | 484.003 | 73.488 |
| 70741 | 20563.777 | 7286.501 | 1143.727 | 588.902 | 3139.349 | 112.588 |
| 59596 | 16428.303 | 16410.354 | 1995.231 | 373.15 | 97.243 | 886.47 |
| 56448 | 8764.309 | 28089.33 | 3578.12 | 289.184 | 22.105 | 9.317 |
| 56420 | 15931.473 | 9161.576 | 13093.543 | 585.337 | 37.682 | 9.127 |
| 57090 | 7559.305 | 30718.529 | 6226.15 | 4887.683 | 283.504 | 18.081 |
| 41920 | 14522.98 | 4873.693 | 6783.85 | 584.118 | 1035.664 | 43.296 |
| 33599 | 9880.994 | 4708.252 | 812.33 | 1086.089 | 65.835 | 152.233 |
| 38465 | 3779.036 | 13497.126 | 3739.924 | 473.079 | 392.189 | 16.481 |
| 38700 | 2222.899 | 3686.353 | 4277.55 | 1081.223 | 273.049 | 118.803 |
| 37208 | 789.787 | 9229.84 | 3128.155 | 1025.456 | 426.614 | 90.387 |
| 36689 | 1146.222 | 1977.49 | 9664.041 | 1183.655 | 229.857 | 68.248 |
| 38080 | 3803.96 | 3110.436 | 1942.945 | 5805.497 | 1181.95 | 138.395 |
| 29561 | 3965.733 | 2170.117 | 1220.296 | 382.107 | 2024.552 | 218.843 |
| 26365 | 18813.885 | 6473.455 | 1248.851 | 327.561 | 171.234 | 557.447 |
| 19960 | 1423.965 | 4902.12 | 1815.778 | 359.211 | 53.845 | 24.911 |
| 26332 | 8664.831 | 3706.126 | 2068.674 | 916.903 | 142.281 | 19.137 |
| 21383 | 7392.194 | 8210.657 | 1658.022 | 1078.674 | 218.449 | 22.005 |
| 39350 | 2182.008 | 1845.431 | 4488.746 | 1282.547 | 272.354 | 186.923 |
| 27664 | 2699.332 | 2964.297 | 687.892 | 940.682 | 279.68 | 34.508 |
| 25787 | 4160.412 | 2318.718 | 3285.513 | 305.785 | 290.789 | 53.282 |
| 20273 | 7513.958 | 5370.645 | 1341.721 | 1622.613 | 102.037 | 101.204 |
| 24315 | 1509.725 | 6046.03 | 2291.531 | 675.422 | 789.292 | 22.916 |
| 21305 | 1725.208 | 3310.909 | 2498.717 | 701.186 | 108.245 | 140.133 |
| 21950 | 721.806 | 2616.333 | 2260.832 | 970.329 | 298.966 | 83.208 |
| 15205 | 1270.19 | 2353.781 | 1371.875 | 819.771 | 297.3 | 67.732 |
| 11449 | 1096.1 | 1273.361 | 1933.262 | 696.409 | 187.498 | 33.748 |
| 11166 | 4251.142 | 1659.104 | 1010.394 | 614.297 | 265.65 | 62.355 |
| 8638 | 823.21 | 2152.386 | 706.708 | 294.599 | 179.097 | 43.194 |
| 6431 | 2601.077 | 887.944 | 755.637 | 152.896 | 66.565 | 19.536 |
| 5893 | 728.924 | 1007.442 | 454.373 | 240.788 | 40.285 | 22.082 |
| 3817 | 335.558 | 583.357 | 482.121 | 132.428 | 40.991 | 2.935 |
| 2370 | 3130.339 | 260.924 | 133.135 | 290.007 | 34.543 | 8.6 |
| 1173 | 7323.289 | 758.611 | 165.379 | 83.46 | 77.222 | 2.096 |
| 476 | 676.408 | 225.196 | 143.246 | 10.154 | 15.355 | 3.048 |

Table B.2. continued.

SCONTR: Scottish Nephrops Trawl: Effort in hours: Numbers-at-age (thousands)

| 1965 | 2005 |  | 1 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0 |  |  |  |
| 1 | 6 |  |  |  |  |
| 101975 | 1659.715 | 453.604 | 1101.02 | 102.448 | 4.875 |
| 116972 | 613.623 | 951.561 | 154.546 | 785.807 | 44.575 |
| 135811 | 1788.967 | 2002.916 | 444.377 | 15.668 | 322.969 |
| 166713 | 1761.346 | 1850.07 | 637.399 | 159.199 | 12.641 |
| 155131 | 736.536 | 2706.572 | 437.098 | 155.072 | 44.263 |
| 144704 | 439.172 | 645.419 | 1379.363 | 127.922 | 31.719 |
| 127638 | 1072.488 | 444.198 | 235.897 | 1405.7 | 60.499 |
| 185397 | 3744.591 | 1908.742 | 232.266 | 70.731 | 730.108 |
| 186342 | 3462.89 | 5445.012 | 486.932 | 168.428 | 24.824 |
| 186342 | 1933.55 | 5427.964 | 650.405 | 87.286 | 11.605 |
| 203053 | 5916.971 | 2730.363 | 2846.712 | 319.449 | 35.425 |
| 224347 | 4061.224 | 4343.339 | 893.637 | 1142.92 | 125.278 |
| 196403 | 3573.612 | 1393.724 | 1431.401 | 168.241 | 289.689 |
| 219562 | 6053.242 | 2596.492 | 417.688 | 570.766 | 110.339 |
| 273713 | 659.614 | 3413.303 | 934.795 | 207.461 | 216.936 |
| 254147 | 1439.22 | 1529.161 | 1377.826 | 281.539 | 44.696 |
| 286461 | 1090.91 | 5250.686 | 1199.303 | 430.934 | 105.108 |
| 288902 | 2882.413 | 422 | 2552.725 | 439.981 | 95.697 |
| 293396 | 2702.936 | 1289.896 | 464.524 | 1258.148 | 205.504 |
| 312947 | 15763.118 | 731.211 | 414.638 | 132.72 | 870.58 |
| 384215 | 14885.186 | 3109.454 | 505.209 | 225.601 | 91.132 |
| 368971 | 2231.072 | 1259.03 | 707.734 | 246.405 | 8.838 |
| 395355 | 12048.819 | 1562.25 | 799.307 | 375.73 | 43.994 |
| 397682 | 19926.506 | 12751.985 | 539.705 | 138.471 | 31.741 |
| 379169 | 9854.602 | 485.161 | 443.582 | 152.424 | 71.883 |
| 390391 | 7434.593 | 1407.942 | 58.831 | 63.502 | 8.758 |
| 414817 | 13745.576 | 1280.079 | 294.651 | 27.112 | 43.958 |
| 391325 | 15245.132 | 3122.017 | 453.21 | 211.635 | 19.575 |
| 406753 | 6063.665 | 2833.312 | 611.27 | 159.111 | 112.856 |
| 380688 | 22785.318 | 4821.332 | 2174.707 | 613.104 | 18.004 |
| 333756 | 14759.284 | 5645.468 | 494.013 | 362.773 | 33.499 |
| 345007 | 14700.369 | 1316.965 | 633.638 | 192.741 | 44.427 |
| 354884 | 7854.017 | 1893.631 | 387.294 | 176.713 | 17.444 |
| 350882 | 13268.769 | 1926.434 | 620.474 | 116.935 | 63.417 |
| 337585 | 7208.116 | 1905.577 | 475.713 | 92.945 | 80.71 |
| 332659 | 31208.406 | 934.503 | 360.23 | 101.447 | 28.855 |
| 305743 | 1743.097 | 1271.809 | 189.3 | 80.436 | 14.844 |
| 258169 | 7281.766 | 1291.392 | 483.271 | 29.948 | 8.517 |
| 255729 | 4468.485 | 586.213 | 191.646 | 197.557 | 41.643 |
| 232356 | 3881.27 | 1310.954 | 239.992 | 157.625 | 102.126 |
| 220936 | 1738.881 | 829.542 | 258.178 | 41.47 | 16.707 |

Table B.2. continued.

| IreOTB : Irish otter trawl - Effort in hours - numbers at age (thousands) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} 1995 \\ \hline 1 \end{gathered}$ | $2005$ |  |  |  |  |  |  |
|  | 1 | 0 | 1 |  |  |  |  |
| 1 | 7 |  |  |  |  |  |  |
| 56335 | 222 | 298 | 530 | 461 | 92 | 28 | 98 |
| 60709 | 165 | 531 | 670 | 281 | 175 | 33 | 12 |
| 62698 | 99 | 358 | 515 | 282 | 339 | 133 | 89 |
| 57403 | 51 | 1092 | 552 | 312 | 186 | 218 | 232 |
| $53192$ | 98 | 315 | 437 | 266 | 198 | 109 | 123 |
| $46913$ | 50 | 131 | 188 | 303 | 158 | 76 | 65 |
| 48358 | 14 | 304 | 144 | 101 | 126 | 100 | 44 |
| 37231 | 31 | 162 | 388 | 27 | 65 | 97 | 47 |
| 39803 | 90 | 294 | 604 | 492 | 131 | 30 | 0 |
| 35140 | 33 | 387 | 266 | 245 | 200 | 28 | 21 |
| 30941 | 23 | 159 | 188 | 78 | 41 | 19 | 2 |



Figure B.1. Map of the west coast of Scotland showing the catch per unit of effort of whiting during the 2009 Scottish fourth quarter west coast groundfish survey. Each circle is centred on the sample location and the size of the circle is proportional to the number density ( $\mathrm{n} / 30 \mathrm{~min}$ fished) of whiting at age $1+$, according to the legend (top left).


Figure B.2. Map of the west coast of Scotland showing the catch per unit of effort of whiting during the 2010 Scottish first quarter west coast groundfish survey (ScoGFSQ1). Each circle is centred on the sample location and the size of the circle is proportional to the number density ( $n / 30$ $\min$ fished) of whiting at age $1+$, according to the legend (top left).

## Stock Annex 3.5: Nephrops in Vla FU1 1

- Stock Annex 3.5 Nephrops VIa FU11: for latest update see WGCSE 2009, Annex 03.5 Nephrops VIa FU11.


## Stock Annex 3.6: Nephrops in VIa FU1 2

- Stock Annex 3.6 Nephrops VIa FU12: for latest update see WGCSE 2009, Annex 03.6 Nephrops VIa FU12.


## Stock Annex 3.7: Nephrops in VIa FU13

- Stock Annex 3.7 Nephrops VIa FU13: for latest update see WGCSE 2009, Annex 03.7 Nephrops VIa FU13.


## Stock Annex 4.3: Haddock in Division VIb

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Haddock in Division VIb |
| :--- | :--- |
| Working Group: | WGCSE |
| Date | 20 May 2010 |
| Revised by | Vladimir Khlivnoy, Andrzej Jaworski |

## A. General

## A.1. Stock definition

The haddock stock at Rockall is an entirely separate stock from that on the continental shelf of the British Isles (Chuksin and Gerber, 1976; Shestov, 1977; Blacker, 1982; Newton et al., 2008). The TAC for haddock VIb was previously (before 2004) set for Subarea Vb, VI, XII and XIV combined, with a limitation on the amount to be taken in Vb and VIa. In 2004, the TAC for Division VI was split and the VIb TAC for haddock was included with Divisions XII and XIV. This combined TAC has been in place since then.

## A.2. Fishery

The development of the Rockall haddock fishery is documented in the 2001 Working Group Rneport (ICES-WGNSDS, 2001) and in the Report of the ICES Group meeting on Rockall haddock convened in January 2001 (ICES, WGNSDS, 2002). That meeting was set up to respond to a NEAFC request for information on the Rockall haddock fishery. NEAFC agreed to consider regulation of the international fishery in 2001.

The Rockall haddock fishery changed markedly in 1999 when a revision of the EU EEZ placed the southwestern part of the Rockall plateau in international waters. This has opened opportunities for other nations, notably Russia, to exploit the fishery in this area. The table of official statistics includes Russian catches from the Rockall area.

The Russian fleet started fishing operations in international waters at Rockall in MayOctober 1999. The Russian haddock fishery uses bottom trawls with cod-end mesh size of $40-100 \mathrm{~mm}$ (mainly $40-70 \mathrm{~mm}$ ) and retains haddock of all length classes in the catch. This fishery targets concentrations of haddock mainly during the spring and the beginning of summer. Russian catches increased from 458 t in 1999 to 2154 t in 2000. In 2001, they were markedly reduced to 630 t due to the introduction of a closed area and low density of fish concentrations. Russian catches increased again in 20022004 from 1630 to 5844 t . In 2005-2007, they decreased from 4708 t to 1282 t , and are estimated to be 1669 t in 2008.

Prior to 1999, the UK and Ireland fisheries had been principally summer fisheries but in more recent years the Scottish and Irish fishery was conducted throughout the year with the peak in April-May. This shift in the fishery appears to have followed the discovery of concentrations of haddock in deeper water to the west of Rockall, at depths between 200 and 400 m . High catch rates attracted effort into the area. However, catch rates in 2000 were reported to be poor in deeper water. Anecdotal evidence suggests that increased discarding has been associated with the deeper-water fishery compared to the traditional fishery at northern Rockall. In 2004-2007, a con-
siderable proportion of EU landings were taken in the international waters. Historical fishing patterns of the Scottish fleet at Rockall are presented by Newton et al. (2004).

There are some indications that, due to a general decline in catches by the Scottish and Irish fleets in Division VIa, there is an increasing focus in the Rockall fishery in Division VIb (ICES, WGFTFB, 2007). Paired gear (both seine and trawl) are to be tested by some Scottish fishermen, which, if it proves successful, can lead to a considerable increase in effective effort in VIb. The fishery at Rockall seems particularly attractive given the lack of effort restrictions in this area.

Information on the Russian fishery and biological investigations from commercial vessels fishing in Rockall during 2008 are presented in WD11 to WGCSE 2009.

An analysis of the spatial and depth distributions of Rockall haddock in association with oceanographic variables is presented by Vinnichenko and Sentyabov (2004), a WD to WGNSDS 2004. Changes in distribution have occurred over a period coincidental with changes in oceanographic variables. Information on oceanographic conditions on Rockall bank in spring 2005 was presented by Sentyabov at WGNSDS 2005.

## A.3. Ecosystem aspects

In May 2001, the International Waters component of statistical rectangle 42D5, which is mainly at depths less than 200 m , was closed by NEAFC to all fishing activities, except with longlines. That area had the following coordinates:

| LATITUDE | LONGITUDE |
| :--- | :--- |
| $57.000^{\circ} \mathrm{N}$ | $15.000^{\circ} \mathrm{W}$ |
| $57.000^{\circ} \mathrm{N}$ | $14.700^{\circ} \mathrm{W}$ |
| $56.575^{\circ} \mathrm{N}$ | $14.327^{\circ} \mathrm{W}$ |
| $56.500^{\circ} \mathrm{N}$ | $14.450^{\circ} \mathrm{W}$ |
| $56.500^{\circ} \mathrm{N}$ | $15.000^{\circ} \mathrm{W}$ |

In spring 2002, the EU component of this rectangle, again mostly shallow water, was also closed to trawling activities (EC No 2287/2003). The whole Rockall Haddock Box is bounded by the following coordinates:

| LATITUDE | LONGITUDE |
| :--- | :--- |
| $57^{\circ} 00^{\prime} \mathrm{N}$ | $15^{\circ} 00^{\prime} \mathrm{W}$ |
| $57^{\circ} 00^{\prime} \mathrm{N}$ | $14^{\circ} 00^{\prime} \mathrm{W}$ |
| $56^{\circ} 30^{\prime} \mathrm{N}$ | $14^{\circ} 00^{\prime} \mathrm{W}$ |
| $56^{\circ} 30^{\prime} \mathrm{N}$ | $15^{\circ} 00^{\prime} \mathrm{W}$ |

At the 25th Annual Meeting of NEAFC (in November 2006), a closure of three areas on the Rockall Bank to bottom fishery was proposed to protect cold-water corals: North West Rockall, Logachev Mounds and West Rockall Mounds (NEAFC AM, 2006). This measure will be in force for the period January 2007-December 2009.

In 2007, the ICES prepared advice for NEAFC and arrived at the conclusion about the expediency of establishing a new closed area on the so-called Empress of British Banks and adjusting the boundaries of the currently closed area of Northwest Rockall. At the 26th Annual Meeting of NEAFC (in November 2007), a new closed area (Empress of British Banks) was established, and the boundaries of the Northwest Rockall closure were slightly modified (NEAFC AM, 2007). Due to the complex shape of the boundaries of the Northwest Rockall closure proposed by ICES, which poten-
tially could cause problems with enforcement, the introduced changes differed from the ICES recommendation. NEAFC also requested ICES to continue providing all available new information on distribution of vulnerable habitats in the NEAFC Convention Area and fisheries activities in and in the vicinity of such habitats.

WGDEC supported the ICES conclusion on the necessity of revising the boundaries of the Northwest Rockall area established to protect cold+water corals and recommended to consider proposals at the WGNSDS meeting. These recent proposals greatly simplify the boundaries, which would create better conditions for enforcement (see WD8 to WGNSDS, 2008).

## B. Data

## B.1. Commercial catch

## Landings

Nominal landings as reported to ICES are given in Table 4.3.1 of the main Report, along with Working Group estimates of total estimated landings. Reported international landings of Rockall haddock in 1991-2005 were about 4000-6000 t, except for 2001-2002, when they decreased down to about 2300-3000 t. In 2006, they were also low at 2760 t , but increased slightly to 3348 in 2007 , and 4221 t in 2008 . Revisions to official catch statistics for previous years are also shown in Table 4.3.1.

Anecdotal evidence suggests that misreporting of haddock from Rockall have occurred historically (which may have led to discrepancies in assessment), but an estimation of overall magnitude is not possible.

Age composition and mean weight-by-age of Scottish and Irish landings were obtained from port sampling. Data on the volume, length-age and weight composition of landings for the period from 1988 to 1998 correspond to values used at this WG.

In 2002, there was no sampling of the Russian catch and therefore the length composition has to be estimated for this year.

In 2002 and 2003, the structure of the Russian fishery on the Rockall Bank was the same: the same vessels were operating with the same gear in the same fishing areas. The relationship between the haddock length composition obtained from the trawl survey and that in the Russian catches is assumed to be the same for 2002 and 2003; i.e. it is assumed that the length dependent selectivity pattern in 2002 is the same as that in 2003 as there no changes to the fishery in these years. The relationship is described as:

$$
\begin{equation*}
P_{L}=S_{L} p_{L} \tag{1}
\end{equation*}
$$

where $P_{L}$ is the proportion of fish with length $L$ in catches, $p_{L}$ is proportion of fish with length $L$ in the stock (survey), and $S_{L}$ is the proportion of fish of length $L$ taken aboard. $S_{L}$ is determined using a theoretical selectivity curve (Stock Annex, Figure 4.3.1) which may be described by the following formula:

$$
\begin{equation*}
S_{L}=\frac{1}{1+\exp \left(S_{1}-S_{2} L\right)} \tag{2}
\end{equation*}
$$

where SL is the proportion of fish of size L taken aboard, L is the size group, S1 and S2 are coefficients.

The selectivity curve (Stock Annex, Figure 4.3.1), fitted to the data on catch measurements in different periods of the Russian fishery in 2003 is described well by Equation 2 with coefficients S1 $=12.539$ and S2 $=0.4951$. The estimated length frequency distributions for 2003 are compared with the measured length frequency distributions for this year in Stock Annex, Figure 4.3.2. The size distribution in the Russian catch in 2002 is then estimated by applying the theoretical selectivity curve to the survey length frequency in 2002.

To determine the age composition in Russian catches in 2002, the combined agelength key for all years of Russian catches was used.

## Discards

The haddock catch estimated by landings is underestimated as a result of unaccounted discarding of small individuals in the Scottish and Irish fisheries in most years. On Russian vessels, the whole catch of haddock is retained onboard and therefore, total catch is equivalent to landings.

Haddock discards onboard Scottish vessels in 1999 and 2001 and Irish vessels in 1995, 1997, 1998, 2000 and 2001 were determined directly. In other years, indirect estimates of discarding were calculated.

The direct estimates from the Scottish trawlers in 1985, 1999 and 2001 showed a higher proportion of discards of small haddock: from 12 to $75 \%$ by weight (Table 4.3.6 in the main report) and up to $80-90 \%$ of catch numbers. Discard trips in 1995, 1997, 1998, 2000 and 2001 showed that discarding by Irish fishing vessels also reaches considerable values (Table 4.3.7 in the main Report).

Total numbers and weight landed and discarded by age on the Scottish observer trips in 1999 and 2001 are presented in Stock Annex, Tables 4.3.1 and 4.3.2.

The analysis of the discard data collected by Scottish scientists in 1999 and 2001 indicated that only a relatively small proportion of fish taken aboard is landed (Figure 4.2.5). The probability of being retained increases with increasing fish length (Stratoudakis et al., 1999; Palsson et al., 2002; Palsson, 2003; Sokolov, 2003). The relationship between the number of individuals caught and number discarded may be described by the following relationship:

$$
\begin{equation*}
N D_{L}=P D_{L} \times N P_{L} \tag{3}
\end{equation*}
$$

where $N D_{L}$ is the number of discarded fish with length $L, N P_{L}$ is the number of fish caught at length $L, P D_{L}$ is the portion of discarded fish at length $L$.

The length composition of fish taken onboard by Scottish and Irish trawlers was calculated by applying the logistic selectivity curve (Stock Annex, Figure 4.3.3) to the haddock stock length composition obtained from the survey. The selectivity parameters were calculated from Scottish and Irish catches taken by trawls with mesh size that are typical for the fleets of those countries operating at Rockall. The parameters were calculated as $S_{1}=12.608$ and $S_{2}=0.4360$ for the Scottish fleet. $S_{1}=26.248$ and $S_{2}=$ 0.8524 were used for Irish catches.

The catch-at-length compositions obtained by the theoretical curve of selectivity agree well with available results of catch measurements in 1999 and 2001and the distributions are compared in Stock Annex, Figure 4.3.4.

The proportion of fish discarded from catches at different sizes may be determined and modelled using a logistic curve (Stock Annex, Figure 4.3.5) described by the following equation:

$$
\begin{equation*}
P D_{L}=\frac{1}{1+\exp \left(-b\left(L-D L_{50}\right)\right)} \tag{4}
\end{equation*}
$$

where $L$ is size group, $D L_{50}$ is the fish length at which $50 \%$ of this size fish caught are discarded and $b$ is a constant reflecting the angle of curve slope. The parameters were determined from research on discards by Scottish vessels (Stock Annex, Table 4.3.3). The following values were used in subsequent calculations: $D L_{50}=34.66 \mathrm{~cm}, b=-$ 0.8764 . The logistic curve of discards may be found using Equation 2 and the coefficient values: $S_{1}=-15.494$ and $S_{2}=-0.4565$.

To determine abundance of discards the following procedure was used:
a ) A theoretical catch-at-length distribution (\%) was calculated by applying the theoretical selectivity curve to the survey length composition.
b ) An estimate of total catch-at-length was made by summing the reported landings-by-length to the number of discards-at-length calculated from the assumed discard ogive and the landings-at-length data.
c ) An intermediate theoretical catch size distribution in numbers is calculated by dividing the estimate of the total numbers retained (numbers greater than 34 cm ) in B by the fraction retained from the theoretical catch length distribution calculated in a).
d ) Theoretical discard size frequency is then calculated by applying the theoretical discard ogive to the intermediate theoretical catch size distribution.

The spreadsheet containing these calculations can be found in the stock file.
Calculations where the discard curve was applied agree well with the results of size composition measurements by Scottish vessels in 1999 and 2001 (Stock Annex, Figure 4.3.6).

Aboard Irish vessels, larger fish are retained (Stock Annex, Figure 4.3.7). The portion of discards was calculated using Equation 2 with coefficients $S_{1}=-10.093$ and $S_{2}=-$ 0.2459, from the combined 1995-2002 Irish discard trips.

The Russian fleet fish in the areas covered only partially by the bottom+trawl surveys. However, Russian vessels retain all haddock and therefore there is no need to calculate discards. There is no information on large-scale fisheries of other countries outside the surveyed area. In addition, available data on the real length composition of catches indicate a correspondence between length composition obtained by the results from surveys and commercial catches, including the catches obtained in the parts of Russian fishery (Stock Annex, Figures 4.3 .2 and 4.3.6).

The amount of discarded haddock by age was determined using a length-age key derived by the data collected during the trawl survey allowing for selectivity of the fishery (Stock Annex, Figure 4.3.3).

In 1998 and 2000, the trawl survey for haddock in the Rockall Bank area was not carried out. To determine the haddock length composition in these years, the length distribution was calculated from the survey data in the previous and following years.

For this purpose, the length-age matrices characterizing the stock status in the years before and after the missing data year were obtained. The length-age distribution
from the year before the missing year was projected forward on the basis of mean growth increment at age and estimated total mortality. Similarly the distribution from the year after was projected backwards. The length composition in the missing year was then calculated from these two estimates.

The total loss $(Z)$ used in the calculation described above was determined by minimization of values of deviation square sum between survey age group abundance values in previous and following years by the data from surveys and calculated data. At that, the factor of age effect $\left(S_{a}\right)$ was taken into account. The mean growth increment at age was also estimated from the survey data. The method of calculation is explained further in WD8 to WGNSD 2004 and a spreadsheet showing the calculations is in the stock file.

## B.2. Biological

Age composition and mean weight-at-age of Scottish and Irish landings were obtained from port sampling.

Age composition and mean weight-at-age of Russian landings were obtained by observers onboard commercial fishing vessels. In 2002, there was no sampling of the Russian catch and therefore the length composition for that year had to be estimated (for estimation details, see Stock Annex). Observer data from commercial vessels are also available for Norwegian landings for 2006-2008.

In the absence of any direct estimates of natural mortality, M has been set at 0.2 for all ages and years.

Natural mortality coefficient and portion of mature individuals by age used for estimation correspond to those adopted by Working Group before.

Previous Working Groups have adopted a maturity ogive with knife-edge maturity-at-age 3 in assessments of this stock (see the Table below).

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $7+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Proportion mature | 0 | 0 | 1 | 1 | 1 | 1 | 1 |

The data from new Russian histological examination of haddock gonad samples mass sexual maturation occurs at age of two years with length of 25 cm (WGNSDS WD6 2006). These data agree well with the results of recent Scottish research in compliance with which the majority of fish become mature at the age of 2 years (ICES 2003; Newton et al., 2004). Visual estimation of maturity stage of post-spawning haddock on the Rockall Bank in expeditions leads to considerable errors. For more precise estimation of length and age-at-maturity for haddock it is necessary to conduct investigations in pre-spawning and spawning periods as well as to collect gonads for further histological analysis (see WGNSDS WD6 2006 for further details).

Research on determining more precise values for natural mortality and maturity ogive parameters should be continued and new estimates could be used in future stock assessments.

In the absence of any direct estimates of natural mortality, M has been set at 0.2 for all ages and years. MSVPA estimates for the North Sea haddock stock give estimates of M of 2.05 at age $0,1.65$ at age $1,0.40$ at age $2,0.25$ at ages 2 and 4 , and 0.20 at ages $5+$ (ICES CM 2003/ACFM:02). Similarly, large values of $M$ at the younger ages at Rockall would have implications for interpretation of fishing mortality patterns from survey-
based methods such as SURBA which essentially estimate total mortality conditional upon assumptions regarding survey catchability-at-age.

ACFM in 2001 encouraged the WG to investigate a more realistic maturity ogive for this stock. At the 2002 Working Group combined sex maturity ogives were presented to the WG for Russian sampling in 2000-2001 and Scottish sampling in 2002. In 2003 new sex disaggregated maturity data were supplied to the Working Group for Russian sampling. The results of all these recent studies indicate that a high proportion of both females and males at age 2 were mature.

## B.3. Surveys

There is only one research survey index available for VPA assessment of this stock from the Scottish survey conducted annually in September (Figure 4.2.4, Table 4.2.8). However, from 1997 onwards the Scottish survey was only conducted in alternate years. Due to concerns about the haddock stock at Rockall some extra time was allocated to carry out a partial survey in September 2002. Full surveys have been conducted since 2005 to improve the quality of assessment. The Scottish survey is currently conducted on about 40 (the target number for a survey) standard trawl stations. However, the survey area and number of stations varied in different years. The majority of stations are within the 200 m depth contour. In 2002 the survey was carried out in the central and northern parts of the bank. In 1999 the survey switched from using an Aberdeen $48^{\prime}$ bottom trawl to a GOV trawl and from 60 min tows to 30 min tows. The indices have been adjusted for tow duration, but no calibration has been made for gear changes. A 20 mm mesh size is used on the survey.

In spring 2005, the Russian trawl-acoustic survey (TAS) for haddock on the Rockall Bank was conducted for the first time (Oganin et al., 2005). However, no such survey has been carried out in subsequent years. In the 2005 survey, the trawl survey method estimated the total stock number at 190.63 million individuals and its biomass at 43400 t (see the Table below). The acoustic survey yielded a haddock biomass estimate of 60000 t with the abundance of 225.9 million (see the WGNSDS 2006 Report for more details of the trawl-acoustic survey). The estimates of haddock abundance and biomass from the two methods are quite similar. The results of the Russian trawl-acoustic survey are summarised in the Table below:

|  |  | Area |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: |
| Survey <br> type | Area |  |
| component |  |  |

* Pelagic component estimated to make up 13.7\%.

The Irish Fisheries Board (BIM) and the Marine Institute recently conducted a collaborative series of surveys to assess the length structure of haddock at various locations on the Rockall Bank and tested the selectivity of a number of codend configurations, which are typically used by both the Irish and Russian fleets.

## B.4. Commercial cpue

Commercial cpue series are available for Scottish trawlers, light trawlers, seiners, Irish otter trawlers and Russian trawlers fishing in VIb. The effort data for these five fleets are shown in Figure 4.2.1 and Table 4.2.7. Commercial cpue series for the different fleets are shown in Figure 4.2.2.

In 2005-2007, the Russian effort in bottom fishery (in hours and number of vessels/days) decreased due to economic reasons. The effort in 2008 increased slightly compared to 2007. Haddock catches varied accordingly with the changes in fishing effort. In 2006-2007, fishing efficiency in the Russian haddock fishery (mainly with trawlers of tonnage class 10) increased compared to previous years. In 2008, with trawlers of class 8 and 9 only, it was still high (on average, 12.2 t per fishing day for trawlers of class 9), but lower than the efficiency in 2007 (on average, 16.9 t per fishing day for a trawler of class 10). In the period of the targeted fishery (April-May), the mean catch of haddock per hour trawling by a trawler of tonnage class 9 was 0.86 t (in 2007, it was 0.88 t for a trawler of class 10) (Figure 4.2.2). The dynamics of catch per unit of effort for this type of vessels agrees well with year-to-year variations in total biomass of haddock (Figure 4.2.3).

The effort data from the Scottish fleets are known to be unreliable due to changes in the practices of effort recording and non-mandatory effort reporting (see the Report of WGNSSK 2000, CM 2001/ACFM:07, for further details). It is unknown what proportion of Scottish and Irish effort was applied directly to the haddock fishery. The apparent effort increase may just be the result of more exact reporting of effort due to VMS, but another suggestion is that it arises from a 'days at sea' measure. Working at Rockall keeps 'days at sea' elsewhere intact (the years in question do correspond to the introduction of the days at sea legislation) and it is possible that vessels are either working extra days in VIb or they are simply reporting extra days from VIb. It is difficult to conclude which of these scenarios is more likely.

The Irish otter trawl effort-series indicated low values between 2002 and 2005 with the lowest value in 2004. In 2006-2008, the effort increased considerably.

The WG decided that the commercial cpue data, which do not include discards and have not been corrected for changes in fishing power despite known changes in vessel size, engine power, fish-finding technology and net design, were unsuitable for catch-at-age tuning.

## B.5. Other relevant data

## C. Historical stock development

Model used:
The assessment is based on catch-at-age data and one survey index (Scottish Groundfish Survey) and conducted using the XSA method.

Software used:
XSA from Lowestoft suite of VPA programs
Model Options chosen:
Settings for the final XSA assessment in the recent years are shown in the Table below.

| Assessment year | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Assessment model | XSA | XSA | XSA | XSA | XSA |
| Time series weights | none | none | none | none | none |
| Model | power | power | power | power | power |
| Catchability dependent for ages < | 4 | 4 | 4 | 4 | 4 |
| Regression type | C | C | C | C | C |
| Q plateau | 5 | 5 | 5 | 5 | 5 |
| Shk se | 1.0 | 1.0 | 1.0 | 1.0 | 1.0 |
| Shk age-yr | 4 yrs | 4 yrs | 4 yrs | 4 yrs | 4 yrs |
|  | 3 ages | 3 ages | 3 ages | 3 ages | 3 ages |
| Min se | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
| Plus group | 7 | 7 | 7 | 7 | 7 |
| Fbar | 2-5 | 2-5 | 2-5 | 2-5 | 2-5 |

Input data types and characteristics:

| Type | Name | Year range | Age range | Variable from <br> year to year <br> Yes $/$ No |
| :--- | :--- | :---: | :---: | :--- |
| Caton | Catch in tonnes | $1991-2008$ | $1-7+$ | Yes |
| Canum | Catch-at-age in <br> numbers | $1991-2008$ | $1-7+$ | Yes |
| Weca | Weight-at-age in <br> the commercial <br> catch | $1991-2008$ | $1-7+$ | Yes |
| West | Weight-at-age of <br> the spawning stock <br> at spawning time. | $1991-2008$ | $1-7+$ | Yes |
| Mprop | Proportion of <br> natural mortality <br> before spawning | $1991-2008$ | $1-7+$ | No, set to 0 for all <br> ages in all years |
| Fprop | Proportion of <br> fishing mortality <br> before spawning | $1991-2008$ | $1-7+$ | No, set to 0 for all <br> ages in all years |
| Matprop | Proportion mature- <br> at-age | $1991-2008$ | $1-7+$ | No, the same <br> ogive for all years |
| Natmor | Natural mortality | $1991-2008$ | $1-7+$ | No, set to 0.2 for <br> all ages in all <br> years |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :---: | :---: |
| Tuning fleet 1 | SCOGFS | $1991-2008$ | $1-6$ |

## D. Short-term projection

Model used: Age-structured

Software used: MFDP prediction with management option table and yield-perrecruit routines. MLA used for probability profiles and sensitivity analysis.

Initial stock size: Taken from XSA for age 1 and older. The recruitment-at-age 1 in 2009 is estimated using RCT3. For forecasting recruitment in 2010 and thereafter, a geometric mean was used for 1991-2006.

Natural mortality: Set to 0.2 for all ages in all years.
Maturity: The same ogive as in the assessment is used for all years.
$F$ and $M$ before spawning: Set to 0 for all ages in all years.
Weight-at-age in the stock: Three-year means (mean weights in the stock are assumed to be the same as catch weights, see below).

Weight-at-age in the catch: Three-year means.
Exploitation pattern: Average of the three last years. Landings F are varied in the management option table.

Intermediate year assumptions: Status quo F.
Stock-recruitment model used: XSA estimate of recruits at age 1 for intermediate year. RCT3 model. used for intermediate year +1 in 2009 and the long-term geometric mean recruitment-at-age 1 is used for forecasting recruitment in 2010 and thereafter.

Procedures used for splitting projected catches: F vectors in each of the last three years of the assessment are multiplied by the proportion landed at age to give partial F for landings. The vectors of partial F are then averaged over the last three years to give the forecast values.

## E. Medium-term projections

Model used: Age structured
Software used: MLA used for Medium-term projections.
Initial stock size: Taken from the XSA for age 1 and older. The recruitment-at-age 1 in 2009 is estimated using RCT3. For forecasting recruitment in 2010 and thereafter, a geometric mean was used for 1991-2006.

Natural mortality: Set to 0.2 for all ages in all years.
Maturity: The same ogive as in the assessment is used for all years.
F and $\mathbf{M}$ before spawning: Set to 0 for all ages in all years.
Weight-at-age in the stock: Three-year means (mean weights in the stock are assumed to be the same as catch weights, see below).

Weight-at-age in the catch: Three-year means.
Exploitation pattern: Average of the three last years.

## Intermediate year assumptions:

Stock-recruitment model used: RCT3 model used for intermediate year +1 in 2009 .
Uncertainty models used:
1 ) Initial stock size:

2 ) Natural mortality:
3 ) Maturity:
4) F and $M$ before spawning:

5 ) Weight-at-age in the stock:
6 ) Weight-at-age in the catch:
7 ) Exploitation pattern:
8 ) Intermediate year assumptions:
9 ) Stock-recruitment model used:

## F. Yield and biomass-per-recruit/long-term projections

Model used: Yield and biomass-per-recruit over a range of F values.
Software used: MLA and "st graf".
Maturity: Fixed maturity ogive as used in the assessment.
$\mathbf{F}$ and $\mathbf{M}$ before spawning: Set to 0 for all ages in all years.
Weight-at-age in the stock: Three-year means (mean weights in the stock are assumed to be the same as catch weights, see below).

Weight-at-age in the catch: Three-year means.

## G. Biological reference points

Biological reference points for this stock are given below:

| Blim: | 6000 t (lowest observed SSB) |
| :--- | :--- |
| B $_{\mathrm{pa}}:$ | $9000 \mathrm{t}($ Bloss $\times 1.4)$ |
| $\mathrm{F}_{\mathrm{pa}}:$ | 0.4 (by analogy with other haddock stocks). |

## H. Other issues

None.

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Table 4.3.1. Scottish landings and raised discards of haddock in 1999 estimates at Rockall from discard observer trips conducted on Scottish vessels.

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total |
| Landing, N <br> $\left({ }^{*} 1000\right)$ | 0 | 0 | 436.9 | 1211.9 | 1069.5 | 849.4 | 1220.6 | 1432.3 | 411.9 | 87.7 | 0.4 | 0 | 1.4 | 6722 |
| Landing, tonnes | 0 | 0 | 135.8 | 432.5 | 420.7 | 383.9 | 646 | 760.7 | 245.5 | 49.6 | 0.5 | 0 | 4.3 | 3079.5 |
| Discards, N <br> $\left({ }^{*} 1000\right)^{1}$ | 22.414420 .815276 .9 | 6844.7 | 2534.8 | 1516 | 734.3 | 219.4 | 39.6 | 0 | 0 | 0 | 0 | 41609.1 |  |  |
| Discards, <br> tonnes ${ }^{1}$ | 1.5 | 2284.1 | 3658.2 | 1936.2 | 799.1 | 515.4 | 248.8 | 86.2 | 17.6 | 0 | 0 | 0 | 0 | 9547.2 |
| Discards, N <br> $\left({ }^{*} 1000\right)^{2}$ | 12.513306 .115895 .97168 .1 | 2588.9 | 1555.7 | 772.5 | 247.9 | 48.6 | 12.2 | 0.7 | 0 | 0 | 41609.2 |  |  |  |
| Discards, <br> tonnes ${ }^{2}$ | 0.3 | 2241.2 | 3791.3 | 2035.1 | 821.7 | 538.7 | 268 | 103.8 | 22.7 | 6.3 | 0.5 | 0 | 0 | 9829.6 |

${ }^{1}$ raised estimates from discard observer trips at Rockall.
${ }^{2}$ estimates obtained from a logistic discard curve for 1999.

Table 4.3.2. Scottish landings and raised discards of haddock in 2001 estimates at Rockall from discard observer trips conducted aboard Scottish commercial vessels.

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total |
| Landing, N (*1000) | 0 | 0 | 326.5 | 489.1 | 132.9 | 774.3 | 326 | 223.9 | 113.5 | 22.4 | 3.8 | 0 | 0 | 2412.3 |
| Landing, tonnes | 0 | 0 | 128.6 | 157 | 82.4 | 262.4 | 125.2 | 90.2 | 59.3 | 19.9 | 3 | 0 | 0 | 928 |
| Discards, N (*1000) ${ }^{1}$ | 3.1 | 6309.9 | 549.7 | 228.4 | 66.3 | 8.1 | 1 | 0.1 | 0.1 | 0.1 | 0 | 0 | 0 | 7166.8 |
| Discards, tonnes ${ }^{1}$ | 0.2 | 967.4 | 126.8 | 58.7 | 17.8 | 2.4 | 0.3 | 0.1 | 0 | 0 | 0 | 0 | 0 | 1173.8 |
| Discards, N (*1000) ${ }^{2}$ | 531 | 5987.3 | 436.2 | 162.6 | 46.9 | 2.9 | 0.5 | 0.1 | 0 | 0 | 0 | 0 | 0 | 7167.6 |
| Discards, tonnes ${ }^{2}$ | 14.3 | 936.2 | 93 | 38.6 | 11.6 | 0.9 | 0.2 | 0.1 | 0 | 0 | 0 | 0 | 0 | 1094.9 |

${ }^{1}$ raised estimates from discard observer trips at Rockall.
${ }^{2}$ estimates from a logistic discard curve for 2001.

Table 4.3.3. Values of $D L_{50}$ by Scottish discard trips in the Rockall area.

| Year |  | DL50 | b |
| :--- | :--- | :--- | :--- |
|  | 1999 | 36.62 | -0.5923 |
|  | 2001 | 31.20 | -0.8238 |
| Theoretical: | 34.66 | -1.2328 |  |



Figure 4.3.1. Theoretical haddock selectivity curve used to estimate the proportion of haddock lifted onboard Russian trawlers.


Figure 4.3.2. Length distribution of haddock in 2003: 1 - by Scottish groundfish survey, $\mathbf{2 a}$ - by commercial Russian trawlers in June, 2b-by commercial Russian trawlers in July, 3-theoreti-cally-derived.


Figure 4.3.3. Theoretical haddock selectivity curve used to estimate the proportion of haddock lifted onboard Scottish trawlers.


Figure 4.3.4. Length distribution of haddock in 1999 and 2001: 1 - by Scottish groundfish survey, 2 - by commercial Scottish trawlers, 3 - theoretically-derived.


Figure 4.3.5. Selectivity curve used to estimate the proportion of discarded haddock in catches Scottish trawlers.


Figure 4.3.6. Length distribution of discarded haddock in catches Scottish trawlers in 1999 and 2001: 1 - research data; 2 - theoretically-derived.


Figure 4.3.7. Length distribution of haddock landings in VI b (Scottish and Irish data).

## Stock Annex 5.2: Northern Shelf Anglerfish

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Anglerfish (Northern Shelf, Division IIIa, Subarea IV and <br> Subarea VI, and Norwegian Sea, Division IIa) |
| :--- | :--- |
| Working Group | Assessment of Northern Shelf Demersal Stocks <br> 17 May 2005 |
| Date | 19 May 2008 |
| Last updated |  |

## A. General

## A.1. Stock definition

Anglerfish occur in a wide range of depths, from quite shallow inshore waters down to at least 1000 m . Small anglerfish occur over most of the northern North Sea and Division VIa, but large fish, the potential spawners, are more rarely caught. Little is known about when and where anglerfishes spawn in northern European waters and consequently stock structure is unclear. This lack of knowledge is due to the unusual spawning habits of anglerfish. The eggs and larvae are pelagic, but whereas most marine fish produce individual free-floating eggs, anglerfish eggs are spawned in a large, buoyant, gelatinous ribbon which may contain more than a million eggs. Due to this strange behaviour, anglerfish eggs and larvae are rarely caught in conventional surveys.

An EU-funded research project entitled 'Distribution and biology of anglerfish and megrim in the waters to the West of Scotland' (Anon, 2001) did however, improve our understanding. A particle tracking model was use to predict the origins of young fish and indicates that post-larval anglerfish may be transported over considerable distances before settling to the seabed (Hislop et al., 2001). Anglerfish in deeper waters to the west of Scotland and at Rockall could therefore be supplying recruits to the western shelf and the North Sea. Furthermore, results of microsatellite DNA analysis carried out as part of this project show no structuring of the anglerfish stock into multiple genetic populations within or among samples from Divisions IVa, Division VIa and Rockall. In fact this project also suggested that anglerfish from further south (Subarea VII) may also be part of the same stock. Fish tagged and released around the Shetland Islands (Division IVa) by Laurenson et al., 2005 have occasionally been recaptured in Subarea V and also Division IIa.

The WGNSDS considered the stock structure on a wider European scale in 2004, and found insufficient evidence to indicate an extension of the stock area northwards to include Division IIa. Anglerfish in IIa is at present treated separately by the Working Group.

## A.2. Fishery

## A.2.1. Northern Shelf anglerfish fisheries

UK vessels account for more than $50 \%$ of the total reported anglerfish landings from the Northern Shelf area. The Danish and Norwegian fleets are the next most important exploiters of this stock in the North Sea while Irish and French vessels take a significant proportion of the landings to the West of Scotland. The fishery for anglerfish in Subarea VI occurs largely in Division VIa with the UK and France being the most important exploiters, followed by Ireland. Landings from Rockall (Division VIb) are
generally less than 1000 t with the UK taking on average around $50 \%$ of the total. In the North Sea, the majority of landings are reported in Division IVa which reflects the northerly distribution of the species within the North Sea (Knijn et al., 1993).

A general description of the anglerfish fisheries of the most important nations taking part in this fishery is given below:

## Scottish (UK) fishery

The Scottish fishery for anglerfish in Division VIa comprises two main fleets targeting mixed roundfish. The Scottish Light Trawl Fleet (SCOLTR) takes around $60 \%$ of landings and the Scottish Heavy Trawl Fleet (SCOTRL) over $20 \%$. Around $10 \%$ of landings are bycatch from the Nephrops trawlers. The development of a directed fishery for anglerfish has led to considerable changes in the way the Scottish fleet operates. Part of this is a change in the distribution of fishing effort; the development of a directed fishery having led to effort shifting away from traditional roundfish fisheries in inshore areas to more offshore areas and deeper waters. The expansion in area and depth range fished has been accompanied by the development of specific trawls and vessels to exploit the stock. There has been an almost linear increase in landings from Division VIa since the start of the directed fishery until 1996 which has been followed more recently by a very severe decline, indicating the previous increase was almost certainly due only to the expansion and increase in efficiency of the fishery. More recent declines in landings (2002-2004) may have been due to restrictive TACs and the decline is not necessarily representative of the actual landings.

The Scottish fleet operating in VIb consists mainly of large otter trawlers (SCOTRL) targeting haddock and anglerfish at Rockall. Their activity is dependent on weather and the availability of haddock quota in VIb.

The Scottish fishery for anglerfish in the North Sea is located in two main areas: on the Shelf Edge to the north and west of Shetland and at the Fladen Ground. It expanded in a similar manner since the 1980s to that operating in Division VIa. The fishery to the north and west of Shetland operates as an extension to that in Division VIa and consists mainly of light trawlers targeting mixed round-fish. The highest reported landings in recent years (to 2007) come from the statistical rectangles around Shetland. The light-trawler fleet accounted for approximately $55 \%$ of Scottish reported landings in this area in 2007. The landings from the fishery at Fladen are lower but still significant (around $15 \%$ of the total) with anglerfish caught as a bycatch in the Nephrops fishery which consists of approximately 200 vessels in 2007. A small component of the landings ( $\sim 10 \%$ in recent years) comes from the gillnet fishery which operates on the shelf edge in the far northwest of Division IVa. A large proportion of the landings in the gillnet fishery are taken by Spanish owned, UK registered vessels.

Ahead of the anglerfish STECF Review Group meeting in 2006 (SGRST-06-03), attempts were made to develop descriptions of the main Scottish anglerfish fisheries which were spatially more relevant to the stock distribution and activity of fishing vessels, rather than by ICES area. The descriptions used data on catch rates from various sources, including research vessel surveys, observer trips on board commercial boats, consultation with skippers and analysis of individual trip records. An 'anglerfish fishery' area was defined as the combined area of high abundance (catchrates) from FRS/industry survey and observer data analysis. A 'Nephrops fishery' area was assumed to cover the Nephrops grounds which are well defined by soft substrate and are described in the appropriate ICES WGs. The areas are mostly separate but
where overlaps occur, these are taken to be part of the anglerfish area. A third area is defined to include all other statistical rectangles.

In the Scottish 'anglerfish' area, large meshed otter trawlers have the largest contribution to the total landings associated with anglerfish. This métier has a mixed species catch composition with haddock being the most important species and anglerfish and cod the next most important. In the Nephrops area the largest overall landings associated with anglerfish come from the $<100 \mathrm{~mm}$ gear category with the dominant species being Nephrops, followed by haddock and anglerfish.
Previous studied have found it difficult to identify a specific anglerfish fishery as catch composition can vary a great deal over a small spatial scale (i.e. less than a statistical rectangle). Further analysis of the main, large mesh trawl operating in the 'anglerfish area' is required to provide a more comprehensive picture of catch composition. This has so far been beyond the scope of the WG.

## Irish fishery

The Irish fleet which takes around $15-20 \%$ of the total Division VIa landings is a light trawl fleet targeting anglerfish, hake, megrim and other gadoids on the Stanton Bank and on the slope northwest of Ireland. This fleet uses a mesh size of 80 mm or greater. Irish Division VIa landings come mainly from the Stanton bank with some landings from Donegal Bay and the slope northwest of Ireland. Since 1996 there has been an increase in the number of vessels using twin rigs in this fleet. There have also been changes to the fleet composition since 2000, with around ten vessels decommissioned and four new vessels joining the fleet. The activity of this fleet is not thought to have been significantly affected by the recent hake and cod recovery plans.

The Irish fleet otter trawl in Division VIb take anglerfish as a bycatch in the haddock fishery on the Rockall Bank. The fleet targeting haddock uses 100 mm mesh and twin rig trawls. Occasionally Irish-Spanish flag vessels target anglerfish, witch and megrim with 80 mm mesh on the slope in VIb. Discarding practices of these vessels are not known although discarding of anglerfish from the fleet targeting haddock in Division VIb is not thought to be significant (Anon, 2001). The fleet composition changed in 2001. Four vessels have recently been decommissioned and two new vessels have joined the fleet that targets haddock. In 2006 and 2007, the effort of the Irish fleet operating at Rockall has increased with the increase in Rockall haddock TAC.

## Danish fishery

According to logbook records, the majority of Danish anglerfish landings are taken in the northeastern North Sea, in the part constituting the Norwegian Deeps, situated in the Norwegian EEZ of the North Sea. Other important fishing areas for anglerfish are the Fladen Ground (also in IVa) and in the Skagerrak (IIIa). More than $80 \%$ of the Danish landings come from ICES Divisions IVa and IIIa. The remaining part is from the most northern part of Division IVb.

The majority of the Danish vessels are taking anglerfish with demersal trawls with over $90 \%$ of these vessels in the size range $20-40 \mathrm{~m}$.

Fishery definitions by gear type and mesh size as currently used by Danish Fisheries Directorate for the North Sea are given in the following text table:

| Fishery/gear | Mesh size, mm |
| :--- | :--- |
| Dem. Trawl | $>=100 \mathrm{~mm}$ |
| Nephrops trawl | $70-99 \mathrm{~mm}$ |
| Shrimp trawl | $33-69 \mathrm{~mm}$ |
| Industrial trawl | $<=32 \mathrm{~mm}$ |
| Beam trawl | $>=80 \mathrm{~mm}$ |

Note that in the North Sea demersal trawls account for more than $90 \%$ of total Danish landings. However, it is necessary to further specify that at present the majority of the Danish catches of anglerfish are taken by fisheries in the Norwegian zone of IVa applying demersal trawls with mesh size $>=120 \mathrm{~mm}$. In 2006, the fishery with demersal trawl in the Norwegian Deeps (in the Norwegian zone) accounted for around $75 \%$ of total Danish landings by all gears from the entire North Sea. In the Skagerrak (IIIa) the two main fisheries taking anglerfish are the (mixed) Nephrops fishery and the demersal trawl fishery. In both areas minor landings are taken in gillnets and as bycatch in fisheries for shrimp (Pandalus).

Information on the species composition of the landings from Danish fisheries taking anglerfish is available from the Danish logbook records and also from the Danish atsea samples from observers on discard trips. Further details can be found in Section 6.2.1 of ICES WGNSDS 2007. Typically anglerfish constitutes less than $15 \%$ by weight of the landings from demersal trawlers fishing in the Norwegian Deeps.

## Norwegian fisheries

A Norwegian directed gillnet fishery ( 360 mm mesh size), targeting large anglerfish, carried out by small vessels in coastal waters in the eastern part of the Northern North Sea started in the early 1990s. These vessels are responsible for around 60-70\% of the total Norwegian landings from this area and they comprise around $6 \%$ of the total landings from Division IVa since 1999. The remaining Norwegian landings in IVa are mostly bycatch in various trawl fisheries. A similar pattern of fishing is found in the Skagerrak (IIIa). The third quarter has in recent years been the most important season for the directed fishery, while the second quarter is apparently most important for other gears.

## Other fisheries

French demersal trawlers also take a considerable proportion of the total landings from this area. The vessels catching anglerfish may be targeting saithe and other demersal species or fishing in deep water for roundnose grenadier, blue ling or orange roughy.

Since the mid-1990s, a deepwater gillnet fishery targeting anglerfish has been conducting a fishery on the continental slopes to the West of the British Isles, North of Shetland, at Rockall and the Hatton Bank. These vessels, though mostly based in Spain are registered in the UK, Germany and other countries outside the EU such as Panama. Gear loss and discarding of damaged catch are thought to be substantial in this fishery. Until now these fisheries have not been well documented or understood and they seem to be largely unregulated, with little or no information on catch composition, discards and a high degree of suspected misreporting. There are currently (2005) around 16 vessels participating in the fishery, 12 UK registered and four German registered.

In response to the concerns with these gillnet fisheries for deep-water sharks and anglerfish in Subarea VI, the EC banned the setting of gillnets in waters greater than

200 m in 2006 (Council Regulation 51/2006). However, this regulation was reviewed in July 2006 and a new regulation put in place which is a permanent ban, but allows a derogation for entangling nets in waters less than 600 m , not exceeding 100 km in total length with a maximum soak time of 72 hours. (EC Regulation No 40/2008 Annex III, article 8). NEAFC have also introduced an indefinite ban. There is also legislation proposed which will extend the ban to other areas including Division IVa.

In addition, the EU has recently funded a ghost net retrieval programme, DEEPCLEAN, (coordinated by the Marine Institute, Ireland) which is due to commence in autumn 2007. The intention of this programme is to a) maximize the recovery of lost or abandoned gillnets and b) to quantify the scale and biological consequences.

## A.2.2. Division lla anglerfish fisheries

In Division IIa most of the anglerfish is caught by small vessels in a directed gillnet fishery close to the coast. The legal mesh size has, since 1995, been 360 mm and maximum 2 days soaking time. Offshore gillnetting, trawls and Danish seines are responsible for the other catches. For the directed gillnet fishery, the area between N $62^{\circ}$ and $\mathrm{N} 64^{\circ}$ has been the most important with maximum catches almost reaching 3000 tonnes in 1993. During recent years the catches have varied between 1000-2000 tonnes. A fishery north of $\mathrm{N} 64^{\circ}$ has developed rapidly, with catches reaching 2400 tonnes in 2007, exceeding the level of catches in the southern part of IIa for the first time. For the other gears, catches have increased from around 100 tonnes in the early 1990s to approximately 300-500 tonnes during the last four years. Very low catch figures are reported from other nations north of $\mathrm{N} 62^{\circ}$.

## A.3. Ecosystem aspects

No information.

## B. Data

## B.1. Commercial catch

## B.1.1. Data compilation

Quarterly length-frequency distribution data were available from Scotland and Ireland for Division VIa and Spain for Subarea VI in the past. A total international catch-at-length distribution for Division VIa was obtained by summing national raised catch-at-length distributions and then raising this distribution to the WG estimates of total international catch from this area. Landings officially reported to ICES were used for countries not supplying estimates directly to the WG. Since 2001, the Scottish market sampling length-weight relationships (given below) have been used to raise the sampled catch-at-length distribution data Working Group estimates of total landings for Division VIa. Length-frequency data availability for VIb has been limited to Scottish and Irish samples.

| Year Range | Formula (L-length in $\mathbf{c m}, \mathrm{W}-$ <br> weight in g) | Source |
| :--- | :--- | :--- |
| $1992-2000$ | $\mathrm{~W}=0.01626 \mathrm{~L} 2.988$ | Coull et. al., 1989 |
| 2001 onwards | $\mathrm{W}=0.0232 \mathrm{~L} 2.828$ | Scottish Market Sampling |

For anglerfish in the North Sea, catch-at-age composition data are available from Scotland for the years 1992 to 2007. In the past the Scottish quarterly age-length keys were applied to the available length-frequency data and non-sampled catches were
attributed to age assuming their length-frequency distributions to be equivalent to the combined sampled distribution.

As a first step in assembling assessment data for the North Sea component of the stock, length compositions from Scottish market sampling have been raised to Working Group estimates of total landings in the past. The Working Group estimate of total landings was assumed equal to the landings obtained by national scientists plus official landings as reported to ICES for those countries not providing landings data to the Working Group. The Scottish market sampling data are only available from 1993 onwards, and even for these years the level of sampling has been relatively low. More recently, additional length samples are available from the Danish and Norwegian fisheries since 2002 including samples from Division IIIa.

Total international catch-at-length distribution data for the whole Northern shelf (Division IIIa, Subarea IV and Subarea VI) have previously been obtained by summing the length distributions from the individual areas and assuming that this distribution is representative of the whole Northern Shelf. This was then raised to Working Group estimates of total landings for the Northern Shelf.

In addition, catch-at-length distribution data are available from the Norwegian directed coastal gillnetting in Division IIa from 1993 to 2007, although there are no data from 1997-2001. There are also catch-at-length distribution data from anglerfish caught as bycatch in the offshore gillnetting and longlining fleets for 2004-2007. No attempts have been made to present raised catch-at-length distribution for anglerfish from Division IIa.

## B.1.2. Commercial catch data quality

For a number of years, anglerfish in Subarea VI, XII, XIV and Division Vb (EU zone) were subjected to a precautionary TAC ( 8600 t ), based on average landings in earlier years. In 2002 the TAC was set at 4770 t and was further reduced to 3180 t in 2003 and 2004. The TAC was increased in 2005 to 4686 t and to 5155 t for 2007. At the WG in 2003, it was highlighted that the reduction off the TAC in 2003 to just two-thirds of that in 2002 would likely imply an increased incentive to misreport landings and increase discarding unless fishing effort was reduced accordingly (Section 6.4.6, ICES WGNSDS 2003). Anecdotal information from the fishery in 2003 to 2005 appeared to suggest that the TAC was particularly restrictive in these years. The official statistics for these years are, therefore, likely to be particularly unrepresentative of actual landings.

The absence of a TAC for Subarea IV prior to 1999 means that before then, landings in excess of the TAC in other areas, were likely to be misreported into the North Sea. In 1999, a precautionary TAC was introduced for North Sea anglerfish, but unfortunately for current and future reporting purposes, the TAC was set in accord with recent catch levels from the North Sea which includes a substantial amount misreported from Subarea VI. The area misreporting practices have thus become institutionalised and the statistical rectangles immediately east of the $4^{\circ} \mathrm{W}$ boundary (E6 squares) have accounted for a disproportionate part of the combined VIa/North Sea catches of anglerfish.

The Working Group historically (prior to 2005) provided estimates of the actual Division VIa landings by adjusting the reported data for Division VIa to include a proportion of the landings declared from Division IVa in the E6 ICES statistical rectangles. The correction has been applied by first estimating a value for the true catch in each E6 square and then allocating the remainder of the catch into VIa squares in proportion to the reported catches in those squares. The 'true' catches in the E6 squares are
estimated by replacing the reported values by the mean of the catches in the adjacent squares to the east and west. This mean is calculated iteratively to account for increases in catches in the VIa squares resulting from reallocation from the E6 squares. Such a re-allocation of catches may still inadvertently include some landings taken legally in Division IVa on the shelf-edge to the west of Shetland, but these are likely to comprise fish within the distribution of the Division VIa stock component. Due to technical problems associated with changes to the Scottish Executive database and lack of landings data provided to the Working Group by some of the major nations exploiting the fishery, WG estimates of the actual Division VIa landings have not been calculated for recent years (2005-2007).

At the 2010 WGCSE, for data in 2009, this procedure was adjusted to reallocate data to the whole of Area VI: i.e. not just VIa but including Rockall (VIb). This was based on information received from Marine Scotland Compliance indicating that some vessels fishing for anglerfish at Rockall are reporting large catches in the E6 squares from the same voyage. The distribution of landings this new scheme produced was more in keeping with the distribution of the stock as indicated from the anglerfish surveys.

## B.2. Biological

Previous assessments of this stock used the natural mortality rate applied to anglerfish in Division VI adopted by an earlier Hake Assessment Working Group of $0.15 \mathrm{yr}^{-}$ ${ }^{1}$. This value is once more adopted for all ages and lengths in the absence of any direct estimates for this stock.

Historically, the catch-at-age analysis of anglerfish in Division VIa used the same maturity ogive as that applied to anglerfish in Subareas VII and VIII by the Working Group on the Assessment of Southern Shelf Demersal Stocks. However, a number of more recent maturity studies based on the VIa stock indicate that maturity does not occur until much later than previously estimated. Afonso-Dias and Hislop, 1996 give a length-maturity ogive for this stock, $50 \%$ maturity at approximately 74 cm in females, and 50 cm in males. However, this study was based on few samples. New information has become available from the EU-funded project (Anon, 2001) which indicates female $50 \%$ maturity at approximately 94 cm and males at 57 cm . The corresponding age-based ogives indicate $50 \%$ maturity at approximately age 9 in females and age 5 in males. This has also been supported by more recent studies by Laurenson et al., 2005.

## B.3. Surveys

In previous length-based assessments of this stock, a recruitment index was used which had been obtained from the Scottish March West Coast survey. The index consists of numbers of anglerfish less than 30 cm caught per hour. However, at more recent meetings of this WG it has been concluded that the traditional groundfish surveys are ineffective at catching anglerfish and do not provide a reliable indication of stock size. As a result of this conclusion, and the urgent requirement for fishery independent data, Marine Scotland Science began a new joint science/industry survey in 2005. This is a targeted anglerfish survey with a scientific design using commercial gear. In 2006, 2007 and 2009 Ireland extended the anglerfish survey to cover the remaining part of VIa (from $54^{\circ} 30^{\prime}$ to $56^{\circ} 39^{\prime}$ ). Further details of the survey including information on design, sampling protocol and gear and vessel are given in Fernandes et al., 2007 and in annual working documents which describe the survey results.

## B.4. Commercial cpue

## B.4.1. Official logbook data

Previous length-based assessments attempted to use effort data to constrain the temporal trend in fishing mortality. Scottish Light Trawl data, disaggregated into an inshore and offshore component, the latter of which is associated with the anglerfish fishery, for both West of Scotland and Shetland (N Sea) were provided to the Working Group. However, these data are no longer considered to be reliable due to nonmandatory recording of hours fished in the logbook data. Further details of the Scottish fleet effort recording problem can be found in the report of the 2000 WGNSSK (ICES, 2001). Since these data are considered unreliable, they are not presented here.

Irish lpue data in terms of hours fished has been presented to the WG for Division VIa and Division VIb for all fleets up to 2006 (shown in Table B.4.1). The measure of kWdays is believed to be a more reliable proxy for effort than hours fished due to reporting issues and these data are presented in the WG report.

Danish landings and effort data (hours fished) from logbook data are also available to the WG for Division IIIa and Division IVa. Although these data are considered to be reliable (in terms of accuracy of reporting), it is not know to what extent they are useful in providing an indicator of stock size due to management regulations in the Norwegian zone (TAC constraints) and technological creep.

No effort data have been made available to the WG for fisheries operating in Division IIa.

## B.4.2. Tallybook data

Analysis of skippers' personal diary information collected in 2004 and 2005 in an attempt to improve knowledge of the state of the stock and of the Scottish anglerfish fishery provided valuable information to ICES (Bailey, et al., 2004) on temporal and spatial trends in catch rate. Following the success of this data collation exercise, ICES advised the process to continue and a more formal scheme was proposed by FRS.

Extensive discussions with the fishing industry during 2005 resulted in FRS implementing the monkfish tallybook project at the start off 2006. The project is part of a long-term approach to providing better information on the monkfish fishery and the state of the stock, and is being operated in conjunction with fishers' organisations (Scottish Fishermen's Federation, Fishermen's Association Limited and Pecheurs de Manche et Atlantique) and the North Atlantic Fisheries College (NAFC) Marine Centre, Shetland. These organisations have been responsible for distributing the tallybooks, co-ordinating the returns and allocating a vessel code before the anonymised tallybook sheets are forwarded to FRS. The tallybooks are filled in on a haul-by-haul basis to give weight caught by size category and information on haul location, duration and depth in a standardized format as well as gear and mesh being used. Additionally information on mature females has been requested. Data are stored in a database at FRS.

So far, the time-series is relatively short, with the first returns from fishing trips at the end of December 2005 and the most recent from March 2008. Initial participation in the scheme was high with returns received from up to 37 vessels with a wide spatial coverage (across Subarea VI, Division IVa, IIa and Vb ) and different target species. Of the 37 vessels which have so far supplied information, two are French and these are operating towards the southern end of the shelf edge in Division VIa northwest of Ireland. The haul depth information collated so far indicates that most of the hauls
are taken in depths between 100 and 400 m although there are a significant number of hauls from depths between 600 and 800 m . The records from the deeper water are largely from the French vessels although it does appear that a number of the Scottish vessels make occasional trips into deeper water. Average catch rates are similar to those previously seen in the diary data and observer data (presented in previous WG reports) and range from around $10 \mathrm{~kg} / \mathrm{hr}$ for boats targeting Nephrops to over 100 $\mathrm{kg} / \mathrm{hr}$ for some whitefish boats.

Analysis of the catch rate data is presented in the WG report and in Dobby et al., 2007.

## B.5. Other relevant data

None.

## C. Historical stock development

Since 2003 the WG has been unable to provide an assessment of anglerfish. This is due to a combination of unreliable commercial data: landings misreporting in some of the main fleets involved in the fishery and uncertain effort data, and poor catchability of anglerfish in traditional research vessel surveys.

Although, the stock status has been classified as uncertain in recent years, TAC increases of $10 \%$ occurred in both the West of Scotland and North Sea areas on the basis of advice from the STECF Review Group meeting (SGRST-06-03) which examined trends in commercial catch rate data and fishery information.

In previous years the stock assessment has been conducted using a length-based model for which the settings are outlined below.

Model used: Catch-at-length analysis (modified CASA-Sullivan et. al., 1990; Dobby, 2002).

Software used: Fortran coded executable-LBAV4_1.
Model Options chosen:
Sex differentiated von Bertalanffy growth, variability distributed according to a beta function. Parameters taken from Scottish anglerfish survey in 2000: $L_{4}(F)=140.5, K(F)=0.117, L_{4}(M)=110.5, K(M)=0.154$.

Fishing mortality in 1993=1.0
Historical equilibrium fishing mortality fitted using mean of historical WG estimates of landings which is approximately 18000 t over 1987-1991.

Logistic exploitation pattern with fitted parameters.
Trend in temporal fishing mortality equal to trend in recent SCOLTR effort data

Total recruitment normally distributed over length classes
Input data types and characteristics:

| Name | Year range | Variable from year to year <br> Yes/No |
| :--- | :--- | :--- |
| Catch in tonnes | 1993-last data year | Yes |
| Catch-at-length in numbers | 1993-last data year | Yes |
| Weight-at-length in the <br> commercial catch | 1993-last data year | Yes/No-2 weight-length <br> relationships: covering 1993- <br> 2000, and 2001 onwards |
| Weight-at-length of the <br> spawning stock at spawning <br> time. | 1993-last data year | Yes/No-assumed to be the <br> same as weight-at-length in the <br> catch |
| Proportion mature-at-length | 1993-last data year | No-the same ogive for all years |
| Natural mortality | 1993-last data year | No-set to 0.15 for all lengths in <br> all years |

Auxiliary data:

| Type | Name | Year range | Size range |
| :--- | :--- | :--- | :--- |
| Recruitment index | Scottish March West <br> Coast survey | 1993 -last data year | $<30 \mathrm{~cm}$ |

## D. Short-term projection

In previous years the short-term forecast has used a length-structured method with settings outlined below.

Model used: Length-structured
Software used: Fortran coded executable LBForecast.exe
Initial stock size: taken from catch-at-length analysis. The long-term geometric mean recruitment is used in all projection years. Natural mortality: Set to 0.15 for all lengths in all years

Maturity: The same ogive as in the assessment is used for all years
Weight-length relationship: as used in the assessment (Scottish Market sampling)

Exploitation pattern: Fixed exploitation-at-length pattern is estimated in the catch-at-length analysis. This is assumed to apply in all further years.

## E. Medium-term projections

No medium-term projections are carried out for this stock.

## F. Yield and biomass-per-recruit/long-term projections

Previous yield and biomass-per-recruit calculations were carried out on the basis of the results of length-based assessments which are no longer carried out.

## G. Biological reference points

Precautionary approach reference points: "ICES considers that there is currently no biological basis for defining $\mathrm{Blim}_{\text {lim }}$ or $\mathrm{Flim}_{\mathrm{lim}}$. ICES proposes that $\mathrm{F}_{35 \% \text { SPR }}=0.30$ be chosen as $\mathrm{F}_{\mathrm{pa}}$. It is considered to be an approximation of Fmsy."

The statement included above first appeared in 1998, but the WG has been unable to find the basis of the derivation of this reference point and considers it no longer appropriate to include it.

## H. Other issues

In previous ('catch-at-length') assessments of this stock, the SSB was always estimated to be at a very low level. The length data have been based on the U.K. landings only (in Subdivisions. IVa and VIa), where very few individuals over 80 cm appear in the catch and therefore the model predicts very few in the population. Since females do not mature until they are over 90 cm in length the SSB is estimated to be very low. The length data from the eastern part of the North Sea (Danish and Norwegian fisheries) for the recent years indicate a higher amount of larger individuals in the catches. Although the Danish and Norwegian landings are small in comparison to the UK landings, the inclusion of the Danish and Norwegian length frequencies in the data used for any future assessment may change the concept of the magnitude of the SSB.

The fact that mature female anglerfish are rarely observed either on scientific surveys or by observers on board commercial vessels supports a very low estimate of spawn-ing-stock biomass, yet there is little evidence of reduction in spatial distribution as fish are still recruiting to relatively inshore areas. It has been hypothesized that females may become pelagic when spawning as they produce a buoyant, gelatinous ribbon of eggs, and would therefore not appear in the catch of trawlers. (Anglerfish have been caught near the surface, Hislop et al., 2000). This would imply different exploitation patterns for males and females: a dome-shaped pattern (decreased exploitation at larger sizes) for females and a logistic pattern for males. It is also not known whether anglerfish are an iteroparous or semelparous species. The latter would also account for the almost complete absence of spawning females in commercial catches or research vessel surveys.

The key features of the species' life history in relation to its exploitation are the location of the main spawning areas, and whether or not there is any systematic migration of younger fish back into the deeper waters to spawn. At present, despite the large increase in catches during the mid 1990s, there is no apparent contraction in distribution; fish are still recruiting to relatively inshore areas such as the Moray Firth in the northern North Sea. The fact that spawning may occur largely in deep water off the edge of the continental shelf may offer the stock some degree of refuge. However, this assumes that the spawning component of the stock is resident in the deep water, and is thus not subject to exploitation. It is not known to what extent this is true, but if such a reservoir exists then the currently used assessment methods which make dynamic pool assumptions about the population are likely to be inappropriate. Nevertheless, it is clear that further expansion of the fishery into deeper water is likely to have a negative effect on the SSB and given the spatial development of the fishery, it cannot be ruled out that the serial depletion of fishing grounds has been occurring. In addition, some life-history characteristics of anglerfish suggest that it may be particularly vulnerable to high exploitation. A detailed discussion of the fishery development and biology can be found in Sections 7.5.4 and 7.5.5 of the 2000 Report of this Working Group (ICES, 2001).

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Table B.4.1. Anglerfish in Subarea VI. Landings, effort and lpue from the Irish OTB fleet.

|  | IR-OTB-4-6 |  |  | IR-TBB-4-6 |  |  | IR-SCC-4-6 |  |  | IR-GN-4-6 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | IV-VI |  |  | IV-VI |  |  | IV-VI |  |  | IV-VI |  |
| Year | Landings (t) | Effort (hr) | lpue (kg/h) | Landings (t) | Effort (hr) | lpue (kg/h) | Landings (t) | Effort (hr) | lpue (kg/h) | Landings (t) | Effort (hr) | lpue (kg/h) |
| 1995 | 769.21 | 66.54 | 11.56 |  | 0.00 |  | 5.70 | 2.65 | 2.15 | 0.87 | 1.57 | 0.55 |
| 1996 | 698.93 | 68.90 | 10.14 | 16.54 | 1.23 | 13.45 | 4.91 | 2.94 | 1.67 | 1.91 | 2.25 | 0.85 |
| 1997 | 680.78 | 72.71 | 9.36 | 2.055 | 1.07 | 1.93 | 7.79 | 3.00 | 2.60 | 3.40 | 1.83 | 1.86 |
| 1998 | 656.23 | 66.40 | 9.88 | 10.381 | 2.36 | 4.41 | 12.72 | 2.95 | 4.32 | 0.95 | 1.22 | 0.77 |
| 1999 | 512.92 | 63.23 | 8.11 | 1.939 | 1.12 | 1.73 | 12.14 | 4.22 | 2.87 | 6.19 | 0.49 | 12.65 |
| 2000 | 471.95 | 63.33 | 7.45 | 0.045 | 0.13 | 0.35 | 4.64 | 3.86 | 1.20 | 0.87 | 0.11 | 7.60 |
| 2001 | 408.46 | 55.99 | 7.30 | 0.12 | 0.12 | 0.98 | 2.95 | 1.31 | 2.26 | 22.23 | 0.43 | 51.69 |
| 2002 | 317.13 | 40.00 | 7.93 |  | 0.00 |  | 5.06 | 1.58 | 3.20 | 4.94 | 0.23 | 21.48 |
| 2003 | 299.17 | 44.44 | 6.73 |  | 0.00 |  | 3.84 | 2.22 | 1.73 | 1.86 | 0.54 | 3.45 |
| 2004 | 197.89 | 37.50 | 5.28 | 0.176 | 0.35 | 0.50 | 2.15 | 0.98 | 2.20 | 2.46 | 0.54 | 4.57 |
| 2005 | 350.33 | 34.79 | 10.07 |  | 0.04 | 0.00 | 1.07 | 0.69 | 1.56 | 0.00 | 0.04 | 0.00 |
| 2006 | 423.39 | 34.62 | 12.23 | 0.12 | 0.07 | 1.71 | 1.18 | 0.49 | 2.40 | 0.02 | 0.24 | 0.07 |

## Stock Annex 6.2: Cod in VIIa

- Stock Annex 6.2 Cod VIIa: for latest update see WGCSE 2009, Annex 06.2 Cod VIIa


## Stock Annex 6.3: Haddock in VIIa

- Stock Annex 6.3 Haddock in VIIa: for latest update see WGCSE 2009 Annex 06.3 Haddock VIIa


## Stock Annex 6.4: Irish Sea East Nephrops (FU14)

Stock specific documentation of standard assessment procedures used by ICES.

Stock Irish Sea East Nephrops (FU14)<br>Working Group Assessment of Northern Shelf Demersal Stocks<br>Date<br>May 2010

## A. General

## A.1. Stock definition

Throughout its distribution, Nephrops is limited to muddy habitat, and requires sediment with a silt \& clay content of between $30-100 \%$ to excavate its burrows, and this means that the distribution of suitable sediment defines the species distribution. Adult Nephrops only undertake very small scale movements (a few 100 m ) but larval transfer may occur between separate mud patches in some areas. In the eastern Irish Sea the Nephrops stock inhabits an area of muddy sediment extending along the Cumbria coast and its fishery contributes to less than $10 \%$ of overall Irish Sea landings. There is little evidence of mixing between the east and west Irish Sea stocks due to the nature of water current movements in the Irish Sea. The two are treated as separate populations since they have differing population characteristics.

## A.2. The fishery

Between 1999 and 2003 the number of vessels fishing for Nephrops in FU14 declined by $40 \%$ to a fleet of around 50 vessels. This was largely due to the reduction in the number of visiting UK vessels and the decommissioning of part of the Northern Irish and local English fleets. Since then the number of vessels fishing the area has returned to around 80 vessels mainly from Northern Ireland. Currently, around 30 of these vessels, between six and 23 m in length, have their 'home' ports in Whitehaven, Maryport and Fleetwood, England. The rest of the fleet is generally made up of larger vessels from Kilkeel or Portavogie, Northern Ireland.

Between 1987 and 2006, landings from FU14 appeared relatively stable, fluctuating around a long-term average of about 550 t . Landings in 2007, however bucked this trend, and are at their highest level since 1978 at 959 t , this is after landings dropped in 2003 to their lowest apparent level since 1974. The 2008 and 2009 figures of 676 and 694 t respectively are lower than 2007 still remains high, above any other figure recorded since 1990. The introduction of the buyers and sellers legislation in 2006 really precludes direct comparison with previous years as reporting levels are considered to have significantly improved since.

Over the last ten years UK vessels have landed, on average, $87 \%$ of the reported annual international landings. ROI vessels increased their share of the landings to $35 \%$ in 2002 but it has since declined to $2 \%$ in 2009. In 2009, most of the landings were made into England with a high proportion of these landings ( $67 \%$ of the directed landings and $62 \%$ of the total landings) being made by visiting Northern Irish vessels. UK Nephrops directed effort has fluctuated around a downward trend since 1993 but has remained relatively stable since 2003 fluctuating around a mean of 13800 hrs . Changes to recording practices will affect interpretation of the scale of this decline but a decline is real.

The changes to the structure and landing practices of the Northern Irish fleet (see above) will have had some impact on this dataseries. From 2002-2004, fewer of the Northern Irish fleet were landing in England. The differences between lpue figures for individual vessels suggest that earlier years may have included less truly directed effort. Reductions in quota between 2002 and 2006 for VIIa cod and plaice may have restricted total effort in FU14 thereby reducing the more casual effort on Nephrops. Further research is needed to better define the directed fishery. From 2003 the main fleets targeting Nephrops include Nephrops directed single-rig and twin-rig otter trawlers operating out of ports in UK (NI), UK (E\&W) and Ireland.

## Regulations

Regulations introduced as part of a revised package of EC Fisheries Technical Conservation measures in 2000 remain in place. This legislation incorporates a system of 'mesh size ranges' for each of which has been identified a list of target species. In effect, nets in the $70-79 \mathrm{~mm}$ mesh size range must have at least $35 \%$ of the list of target species (which includes Nephrops) and the $80-99 \mathrm{~mm}$ mesh size range requires at least $30 \%$ of the list of target species. A square mesh panel (SMP) of 80 mm is required for $70-79 \mathrm{~mm}$ nets in the Irish Sea. Vessels using twin-rig gear in the Irish Sea must comply with a minimum mesh size of 80 mm (no SMP is required for nets with 80 mm meshes and above).

Other regulations restricting trawling in other fisheries within the Irish Sea will affect effort on these and other stocks. This could either attract local effort or even relocate effort to fisheries in other areas. Although unrestrictive the result of better catch information through the buyers and sellers legislation introduced to the UK from 2006 will have the same effect as quota uptake of stocks which used to be misreported will be quicker.

As well as an Area VII TAC other Nephrops conservation measures in the Irish Sea are a minimum landing size of 20 mm CL length (equivalent to 37 mm tail length or 70 mm total length).

In addition to Nephrops measures the cod spawning areas of the Irish Sea are closed to whitefish directed vessels between 14th February to 30th April part of the Irish Sea cod recovery plan. There is derogation for Nephrops vessels during this closure.

## A.3. Ecosystem aspects

The Working Group has collated no information on the ecosystem aspects of this stock.

## B. Data

## B.1. Commercial catch

Length and sex compositions of Nephrops landed from the Irish Sea East are estimated from port sampling by England and Wales. Length data from this sampling are applied to catch samples collected at sea and raised to total international landings. Catch-length samples are collected independently of landings-length samples but both are considered representative. The independent raising process means that the final annual catch-length frequency distribution still requires scaling to the reported landings. Using a discard ogive derived from samples collected in the early 1990s an initial estimate of discards is taken from the catch distribution. These are then added to the landings distribution to create a dummy catch distribution. The difference between the numbers-at-length for both the raised sampled and dummy catch distribu-
tion was then used to tune a raising factor by minimizing the sums of squares. Once the raising factor is derived, the final discard-length distribution is the difference between the raised catch distribution and the landings distribution and a final catch distribution is a sum of the landings and discard distributions. In 2008 a new discard ogive was calculated from the discard samples collected from 2003 until March 2008 and applied to the 2003 data to date. The lack of discard and catch data between 1995 and 1999 is likely to adversely affect the quality of any analytical assessments. Apparent differences between catch LFDs and discard practices in 1992 to 1994 and 1999 to 2000 are discussed in the Section 5.12 of the 2001 WGNEPH report (ICES, 2001a). 2001 and 2002 catch and landings sampling provided catch compositions to help estimate the LFDs for the missing years. Quarterly discard distributions for the years 1995 to 1999 were estimated by using the discard LFDs for the two preceding and the two following years.

Trial XSAs using these data were attempted at the 2003 WGNEPH. In the absence of routine methods of direct age determination in Nephrops, age compositions of removals were inferred from length compositions by means of 'slicing'. This procedure, introduced at the 1991 WG, uses von Bertalanffy growth parameters to determine length boundaries between age classes. All animals in length classes between boundaries are assigned deterministically to the same age class. The method was implemented in the L2AGE programme which automatically generated the VPA input files. The programme was modified in 1992 to accommodate the two-stage growth pattern of female Nephrops (ICES, 1992) and again in 2001 to separate 'true' as opposed to 'nominal' age classes (ICES, 2001a). The age classes are 'true' to the extent that the first slicing boundary, i.e. lower length boundary for 'age' 0 , is the length-atage zero rather than the lowest length in the data. This was to ensure comparability of 'age' classes across stocks.

## B.2. Biological

Mean weights-at-age for this stock are estimated from studies by Bailey and Chapman, 1983.

A natural mortality rate of 0.3 was assumed for all age classes and years for males and immature females, with a value of 0.2 for mature females. The lower value for mature females reflects the reduced burrow emergence while ovigerous and hence an assumed reduction in predation.

The time-invariant values used for proportion mature-at-age are: males age $1+$ : $100 \%$; females age 1: $0 \%$; age $2+: 100 \%$. The source of these values is not known.

Proportion of F and M prior to spawning was specified as zero to give estimates of spawning-stock biomass at January 1. In the absence of independent estimates, the mean weights-at-age in the total catch were assumed to represent the mean weights in the stock.

## B.3. Surveys

ACFM recommended that UWTV surveys could provide useful fishery-independent data on the status of Nephrops stocks. The UWTV surveys conducted in August 2007 and 2008 are presented here as a preliminary to future assessments. Two previous UWTV surveys were conducted for this fishery in 1997 and 1998 with limited success, because of weather. These surveys and their design were documented at WKNEPHTV (ICES, 2007). The surveys in 2007 and 2008 are consistent but follow a different design to the earlier surveys. For ease of comparison, and consistency, the survey has been based on the current ROI and NI survey in the Western Irish Sea. A
randomised fixed grid ( $3.4 \times 3.4 \mathrm{~nm}$ ) of 34 stations plus a transect of 3 stations in Wigtown bay were sampled. Figure B.3.1 shows the distribution of stations in the TV surveys with the size of the symbol reflecting the Nephrops burrow density.

The survey protocols used were the same, and followed the standards set by WKNEPHTV (ICES, 2007). In 2007 poor visibility hampered the survey and despite repeated attempts at over 15 stations, turbidity scores precluded the use of some of the counts. On first analysis only 20 were considered usable. The 2008 and 2009 survey was far more successful, sea conditions were far better and the quality of the video data collected was much improved. 35 and 32 stations respectively were considered useable. Table B.3.1 provides the estimates for the burrow density and abundance.

These are the first two of a planned series of surveys. Because of uncertainties about the limits of the stock and characteristics of this fishery and in light of SGSURV and WKNEPH (2009) the data will require further analysis and a further survey to qualify the precision of these estimates. These results therefore are only presented as provisional.


Figure B.3.1. Station distribution and relative burrow density, from August TV surveys 2007 to 2009.

Table B.3.1. Irish Sea East (FU14): Results from NI UWTV survey of Nephrops ground.

| Year | Area | No. stations | Non Zero stations | Mean burrow density | Abundance |  |
| :--- | :---: | :---: | :---: | :--- | :---: | :---: |
|  | $\mathrm{km}^{2}$ |  | no. $/ \mathrm{m}^{2}$ |  |  |  |
| $2007^{*}$ | 1043 | 20 | 18 | 0.38 | millions |  |
| $2008^{*}$ | 1043 | 35 | 31 | 0.36 | 393 |  |
| $2009^{*}$ | 1043 | 32 | 28 | 0.25 | 334 |  |

* provisional

A number of factors are suspected to contribute bias to the surveys. In order to use the survey abundance estimate as an absolute it is necessary to correct for these potential biases. The history of bias estimates are given in the following table and are
based on expert opinion on those used in adjacent survey areas which used simulation models, and preliminary experimentation. The biases associated with the estimates of Nephrops abundance in the E. Irish Sea are:

|  | Edge |  | Detection Species |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Time period effect | rate | Cumulative <br> identification |  | Occupancy | bias |
| FU14: Irish Sea East | $<=2009$ | 1.3 | 0.75 | 1.15 | 1 | 1.2 |

Edge effect: Same sledge and set up as Western Irish Sea. Larger burrows systems increase the edge effect.

Detection rate: Same sledge and set up as Western Irish Sea and same staff so detection rate maintained.

Species identification: Factor kept the same as Eastern Irish Sea; Calocaris spp not a perceived problem on Eastern Irish Sea grounds but Goneplax spp. are prevalent across the ground.

## B.5. Other relevant data

When carrying out the XSA in 2003 the landings per unit of effort time-series for the following fleet was used:

England and Wales Nephrops trawl gears. Landings-at-age and effort data from this fishery are used to generate a cpue index. There is also a cpue series from 1995 for Republic of Ireland vessels. Catch-at-age are estimated by raising length sampling of discards and landings to officially recorded landings and slicing into ages (knife-edge slicing using growth parameters). Cpue is estimated using officially recorded effort (hours fished) although the recording of effort is not mandatory. Combined effort for Nephrops trawlers is raised to landings. Discard sampling commenced in 1992 for this fishery, though some years have been missed as discussed above. There is no account taken of any technological creep in the fleet.

## C. Historical stock development

D. Short-term projection
E. Medium-term projections
F. Yield and biomass-per-recruit/long-term projections
G. Biological reference points
H. Other issues

## I. References

Biological Input Parameters

| Parameter | Value | Source |
| :--- | :---: | :--- |
| Discard Survival | 0.00 |  |
| MALES |  |  |
| Growth - K | 0.160 | Irish Sea West data ; Bailey and Chapman (1983) |
| Growth - L(inf) | 60 | $"$ |
| Natural mortality - M | 0.3 | Brander and Bennett (1986, 1989) |
| Length/weight - a | 0.00022 | Hossein et al. (1987) |
| Length/weight - b | 3.348 | $"$ |
| FEMALES |  |  |
| Immature Growth | 0.160 | Irish Sea West data ; Bailey and Chapman (1983) |
| Growth - K | 60 | $"$ |
| Growth - L(inf) | 0.3 | Brander and Bennett (1986, 1989) |
| Natural mortality - M | 24 | Briggs (1988) |
| Size at maturity |  |  |
| Mature Growth | 0.100 | Irish Sea West data ; Bailey and Chapman (1983) |
| Growth - K | 56 | $"$ |
| Growth - L(inf) | 0.2 | Brander and Bennett (1986, 1989) |
| Natural mortality - M | 0.00114 | Hossein et al. (1987) |
| Length/weight - a | 2.820 | $"$ |
| Length/weight - b |  |  |

## Stock Annex 6.5: Irish Sea West Nephrops (FU15)

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Irish Sea West Nephrops (FU15) |
| :--- | :--- |
| Working Group | WKNEPH 2009 (WKNEPH2009) |
| Date | 6 March 2009 |

## A. General

## A.1. Stock definition

Throughout its distribution, Nephrops is limited to muddy habitat, and requires sediment with a silt \& clay content of between $10-100 \%$ to excavate its burrows, and this means that the distribution of suitable sediment defines the species distribution. Adult Nephrops only undertake very small scale movements (a few 100 m ) but larval transfer may occur between separate mud patches in some areas. In the western Irish Sea the Nephrops stock inhabits an extensive area of muddy sediment between the Isle of Man and Northern Ireland and its fishery contributes to more than $90 \%$ of overall Irish Sea landings. There is little evidence of mixing between the east and west Irish Sea stocks due to the nature of water current movements, which is characterised in the west by a gyre, which has a retention affect on both sediment and larvae. The eastern and western Nephrops stocks are treated as separate populations as they have different population characteristics.

## A.3. Ecosystem aspects

A number of studies have examined Nephrops larvae distribution in order to examine how recruitment may impinge upon the distribution of a "catchable" (adult) Nephrops population and the maintenance of the population. Hillis (1968) found that although generally the larvae occupied the same areas as the adults, there was some evidence of advective losses to the southeastern part of their range, most probably due to tidal currents (White et al., 1988). More recent studies in the western Irish Sea have uncovered the existence of a seasonal cyclonic gyre which appears to facilitate retention of larvae over the mud patch (Dickey-Collas et al., 1996; Hill et al., 1996; Horsburgh et al., 2000).

## B. Data

## B.1. Commercial catch

Length and sex compositions of Nephrops landed from the Irish Sea West are estimated from port sampling by Ireland and Northern Ireland and Ireland. A lack of cooperation by the Northern Ireland industry prevented sampling commercial catches over the period 2003-2007. The Irish LFDs are therefore raised to the international catch for these years. Northern Ireland sampling resumed in 2008 and these data are combined with those from Ireland for that year. Sample data is used to compute international removals (Landings + dead discards).

Landings per unit of effort time-series are available from the following fleets:
Northern Ireland Nephrops trawl gears. Landings-at-age and effort data from this fishery since 1986 are used to generate a cpue index. There is also a cpue series since

1995 for a subset of Republic of Ireland Nephrops vessels. Catch-at-age are estimated by raising length sampling of discards and landings to officially recorded landings and slicing into ages (knife-edge slicing using growth parameters). Cpue is estimated using officially recorded effort (hours fished). Discard sampling commenced in the mid-1980s by Northern Ireland and the Republic of Ireland. There is no account taken of any technological creep in the fleet.

## B.2. Biological

Mean weights-at-length for this stock are estimated from studies by Pope and Thomas (1955).

A natural mortality rate of 0.3 was assumed for males and immature females, with a value of 0.2 for mature females. The lower value for mature females reflects the reduced burrow emergence while ovigerous and hence an assumed reduction in predation.

Maturity for females is taken as 22.1 mm carapace length (McQuaid et al., 2006).
Proportion of F and M prior to spawning was specified as zero to give estimates of spawning-stock biomass at January 1. In the absence of independent estimates, the mean weights-at-age in the total catch were assumed to represent the mean weights in the stock.

## B.3. Surveys

Ireland and Northern Ireland jointly carry out underwater television (UWTV) surveys on the main Nephrops grounds in the western Irish Sea (Figure 1) since 2003. These surveys are based on a randomised fixed grid design. The methods used during the survey are similar to those employed for UWTV surveys of Nephrops stocks elsewhere and are detailed in WKNEPHTV, 2007 and WKNEPHBID, 2008.

Northern Ireland have carried out a spring (April) and summer (August) Nephrops trawl surveys since 1994. These surveys provide data on catch rates and length frequency distributions from of stations throughout in the western Irish Sea. These surveys generate data on Nephrops size composition, mean size, maturity and sex ratio.

A number of factors are suspected to contribute bias to the UWTV surveys. In order to use the survey abundance estimate as an absolute it is necessary to correct for these potential biases. The history of bias estimates are given in the following table and are based on simulation models, preliminary experimentation and expert opinion, the biases associated with the estimates of Nephrops abundance in the Irish Sea West are:

|  |  | Edge <br>  <br>  <br> Time period effect | detection <br> rate | species <br> identification occupancy bias |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| FU15: Irish Sea West | $<=2009$ | 1.24 | 0.75 | 1.15 | 1 | 1.14 |

## B.4. Commercial cpue

## B.5. Other relevant data

Table 1 is a summary of available data along with an assessment of its reliability.
Table 2 is a summary of assessment parameters.

## C. Historical stock development

1 ) Survey indices are worked up annually resulting in the TV index.
2 ) Adjust index for bias (see Section B3). The combined effect of these biases is to be applied to the new survey index.
3 ) Generate mean weight in landings. Check the time-series of mean landing weights for evidence of a trend in the most recent period. If there is no firm evidence of a recent trend in mean weight use the average of the three most recent years. If, however, there is strong evidence of a recent trend then apply most recent value (don't attempt to extrapolate the trend further in the future).

## D. Short-term projection

1 ) The catch option table will include the harvest ratios associated with fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max. }}$. These values have been estimated by the Benchmark Workshop (see Section 9.2) and are to be revisited by subsequent benchmark groups. The values are FU specific and have been put in the Stock Annexes.
2 ) Create catch option table on the basis of a range of harvest ratios ranging from 0 to the maximum observed ratio or the ratio equating to $\mathrm{F}_{\text {max, }}$ whichever is the larger. Insert the harvest ratios from step 4 and also the current harvest ratio.
3 ) Multiply the survey index by the harvest ratios to give the number of total removals.
4 ) Create a landings number by applying a discard factor. This conversion factor has been estimated by the Benchmark Workshop and is to be revisited at subsequent benchmark groups. The value is FU specific and has been put in the Stock Annex.
5 ) Produce landings biomass by applying mean weight.
The suggested catch option table format is as follows.

|  | Implied fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Harvest rate | Survey Index | Retained number | Landings (tonnes) |
|  | 0\% | 12345 | 0 | 0.00 |
|  | 2\% | " | 247 | 123.45 |
|  | 4\% | " | 494 | 246.90 |
|  | 6\% | " | 741 | 370.35 |
|  | 8\% | " | 988 | 493.80 |
| $\mathrm{F}_{0.1}$ | 8.60\% | " | 1062 | 530.84 |
|  | 10\% | " | 1235 | 617.25 |
|  | 12\% | " | 1481 | 740.70 |
| $F_{\max }$ | 13.50\% | " | 1667 | 833.29 |
|  | 14\% | ${ }^{\prime}$ | 1728 | 864.15 |
|  | 16\% | " | 1975 | 987.60 |
|  | 18\% | " | 2222 | 1111.05 |
|  | 20\% | " | 2469 | 1234.50 |
|  | 22\% | " | 2716 | 1357.95 |
| $\mathrm{F}_{\text {current }}$ | 21.5\% | " | 2654 | 1327.09 |

## E. Medium-term projections

None presented.

## F. Long-term projections

None presented.

## G. Biological reference points

Harvest ratios equating to fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ were calculated in WKNeph (2009). These calculations assume that the TV survey has a knife-edge selectivity at 17 mm and that the supplied length frequencies represented the population in equilibrium.

$$
\begin{aligned}
& F_{0.1}=10.9 \% \\
& F_{\max }=20.2 \%
\end{aligned}
$$

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Table 1. Summary table of available data.

| FU15 Irish Sea West: Data Available |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Commercial Data | pre-1995 | 1994 | 1995 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| Landings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Effort |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| cpue/lpue |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean size |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sex ratio |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| LFDs |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Catch |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Landings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Discards |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Survey Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Trawl surveys |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Catch rate |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| mean size |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| LFDs |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sex ratio |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Camera Surveys |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Density estimate |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Data Quality |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Poor |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Acceptable |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Reliable |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 2: Biological Input Parameters.

| Parameter | Value | Source |
| :--- | :---: | :--- |
| Discard Survival | 0.10 | ICES (1991a) |
| Discard rate | $40.2 \%$ | 2007 discard sampling. |
| MALES |  |  |
| Growth - K | 0.160 | Hillis (1979) ; ICES (1991a) |
| Growth - L(inf) | 60 | $"$ |
| Natural mortality - M | 0.3 | Brander and Bennett (1986, 1989) |
| Length/weight - a | 0.00032 | After Pope and Thomas (1955) (data for Scottish stocks) |
| Length/weight - b | 3.210 | $"$ |
| FEMALES |  |  |
| Immature Growth |  |  |
| Growth - K | 0.160 | Hillis (1979) ; ICES (1991a) |
| Growth - L(inf) | 60 | $"$ |
| Natural mortality - M | 0.3 | Brander and Bennett (1986, 1989) |
| Size at maturity | 22.1 | McQuaid et al., 2006 |
| Mature Growth |  |  |
| Growth - K | 0.100 | Hillis (1979) ; ICES (1991a) |
| Growth - L(inf) | 56 | $"$ |
| Natural mortality - M | 0.2 | Brander and Bennett (1986, 1989) |
| Length/weight - a | 0.00068 | After Pope and Thomas (1955) (data for Scottish stocks) |
| Length/weight - b | 2.960 | $"$ |



Figure 1: Western Irish Sea Nephrops stations

## Stock Annex 6.6: WhitingVIIa

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Irish Sea Whiting (Division VIIa) |
| :--- | :--- |
| Working Group | Assessment of Northern Shelf Demersal Stocks |
| Last updated | WGNSDS 2008 |
| Updates | Inclusion of Fishery Data from Ireland |

## A. General

## A.1. Stock definition

Whiting in Division VIIa are considered a single stock for management purposes. In 2004 an informal meeting was established to review current knowledge of the distribution, movements and stock structure of whiting in the Irish Sea, and linkages between whiting in the Irish Sea and surrounding management areas. Information on egg and larval, tagging, survey studies was presented as a working document (WD10) in WGNSDS, 2005. The results of this are synopsized below:

UK egg and larva surveys have shown that whiting spawn in spring throughout the eastern Irish Sea and in the coastal waters of the western Irish Sea. This is supported by the distribution of actively spawning fish caught during trawl surveys in March.

Transport of whiting eggs, larvae or pelagic pre-recruits from Celtic Sea spawning grounds into the Irish Sea is likely to be impeded by the Celtic Sea thermal front that becomes increasingly established from spring onwards.

Whiting recruitment grounds are in the same general area as the spawning grounds, and young whiting are widespread in the coastal bights of the Irish Sea. The gyre system that becomes established from late spring onwards in the western Irish Sea appears important in retaining larvae and pelagic pre-recruits of whiting, as shown by the results of frametrawl surveys of pelagic pre-recruits in the western Irish Sea.

As the whiting become demersal from late summer onwards, they are found throughout the western Irish Sea although densities appear highest around the periphery of the mud patch in coastal waters and along the southern boundary between Ireland and the Isle of Man. This pattern is also noted by fishermen operating in this area. Densities of young whiting in the eastern Irish Sea appear highest off Cumbria and the Solway Firth in autumn, but are more widespread in spring.

Tagging studies in the late 1950s show some seasonal dispersal of whiting from the Irish Coast to as far as the Clyde, Liverpool Bay and the Celtic Sea, with evidence of return migrations. Whiting tagged in these studies ranged from about $20-40 \mathrm{~cm}$, averaging around 30 cm . Whiting recaptured well away from the tagging sites off County Down in the western Irish Sea tended to be several cm larger, on average, than the tagged whiting.

Both the western Irish Sea and the Clyde have historically been characterised bycatches of immature and first-maturing whiting, whilst the eastern Irish Sea has a broader agerange of whiting. This pattern persists to the present day.

The evidence for interchange of whiting between the western Irish Sea and other areas within the Irish Sea precludes treating different areas within the Irish Sea as containing functionally separate stocks. Spatial modelling of the populations would require information on rates of dispersal between areas.

Trawl surveys continue to show that juvenile whiting are very abundant in the coastal waters of the Irish Sea, and that whiting are one of the most abundant fish species taken in the surveys. Hence, there have been no indications of depressed recruitment associated with the apparent steep decline in abundance of large whiting. Length at $50 \%$ maturity in female whiting is only $20-21 \mathrm{~cm}$ in the Irish Sea and neighbouring management areas, and spawning appears predominantly by young whiting of 1-3 years old.

## A.2. The fishery

Most landings by the Irish and UK (NI) fleet, which take the bulk of the Division VIIa whiting catch, are from the western Irish Sea (ICES CM 2003/ACFM:04) and are made predominately by single- and twin-rig trawlers. A small number of UK pair trawlers also fish for whiting. The UK (E\&W) fleet has declined substantially over time, and the bulk of its landings are from inshore otter trawlers targeting mixed flatfish and roundfish in the eastern Irish Sea. Discarding in this stock is thought to be high in all fleets, particularly in the Nephrops fishery. The Nephrops directed fishery operates on the main whiting nursery areas in the western Irish Sea, and is particularly intensive in the summer months. The mesh size mainly in use in the fishery is 70 mm in single trawls and 80 mm in twin trawls targeting Nephrops. The western Irish Sea fishery for whiting has declined substantially in recent years, and the increase in abundance of haddock has resulted in few vessels targeting whiting.

Vessels operating with 70 mm and 80 mm mesh are required to use square mesh panels. Square mesh panels were introduced as a technical measure to reduce fishing mortality on whiting. Square mesh panels have been mandatory for all UK trawlers (excluding beam trawlers) in the Irish Sea since 1993 and for Irish trawlers since 1994. While the effects of this technical measure have not been formally evaluated, the Nephrops fishery still generates substantial quantities of whiting discards. Effort by Irish Nephrops trawlers in the main areas of whiting bycatch has shown some reduction during the period of the Irish Sea cod recovery plan closures. However, the summer peak in activity of the Nephrops fishery was not affected by the recovery plans. As the activities of the Nephrops fleet were not restricted by the cod recovery plan, it is unlikely that the recovery plan was effective in reducing levels of discarding in this stock.

There has been some recent decommissioning of vessels in the Irish Sea. Most recently, Ireland introduced a further decommissioning scheme in 2008, which aims to remove 11140 GT from the fleet register. This is targeted at vessels over 10 years of age and $>18 \mathrm{~m}$ in length. To date the majority of applications emanate from east and west coast ports from vessels, which traditionally target Nephrops with uptake from the Southeast also. It is expected that much of the actual effort removed from the decommissioning scheme may be partially negated through the introduction of $\sim 21$ modern second-hand vessels (mostly ex-French) into the fleet over the last few years.

The reported landings of whiting in 1999-2001 by UK vessels decommissioned in 2002 amounted to about $7 \%$ of the total international landings of whiting in those years. Whilst few new Irish vessels have joined the fishery, some vessels from County Donegal
have reported catches of whiting in VIIa. These vessels have been attracted into the Celtic Sea fishery in recent years in response to poor catches in other areas. Irish landings of whiting in the southwestern part of VIIa now contribute the bulk of the total Irish landings in the Division (ICES CM 2003/ACFM:04). The difference in grounds in the southern part of VIIa means that whiting in the area are more likely to function as part of the Celtic Sea stock rather than the Irish Sea stock.

Irish otter board trawlers fishing ICES Area VIIa generally use twin-rig gear to fish for Nephrops. However there are also localized mixed fisheries both in the north and south ends of VIIa. The Irish Sea Nephrops fleet is highly opportunistic and of this fleet, there are only a handful of boats that fish the Irish Sea Prawn Grounds $100 \%$ of the time. The rest of the fleet divides its time between the Irish Sea, Smalls, Aran and Porcupine Grounds dependant on tides, weather and market forces. Because of the need to fish further away from their home port and in rougher sea conditions, many of the older and smaller wooden vessels are being replaced with new and second hand steel vessels. Most of these newer vessels are French-style twin-riggers. To maximize the return on their investment, many of the owners of newer vessels are opting for relief skippers and crews so that the vessels are fishing as much as possible.

In 2006, for the Irish fleet for the first time, Nephrops landings from the Smalls grounds (VIIg) have surpassed those from the Irish Sea grounds. This reflects the increasing amount of effort by East Coast vessels in 7 g where in general, better prices are obtained for their catch. Two significant fleet movements occurred in 2006 for the Irish fleets. Firstly, there was a brief shift in effort by the Nephrops fleet towards the Aran Grounds around October due to reports of good fishing in the area. Also, some of the larger twinriggers in the fleet switched to tuna fishing in the Bay of Biscay during the summer months.

The main species targeted by the otter trawl fleet are Nephrops, cod, ray, haddock, anglerfish and whiting. The Irish beam trawl fleet predominantly targets black sole and other high-quality flatfish and divides its effort between VIIa and VIIg depending on weather, tides and market forces.

For the UK NI fleet decommissioning at the end of 2003 removed 19 out of 237 UK vessels that operated in the Irish Sea, representing a loss of $8 \%$ of the fleet by number and $9.3 \%$ by tonnage. Of these vessels, 13 were vessels that used demersal trawls with mesh size $>=100 \mathrm{~mm}$. The previous round of decommissioning in 2001 removed 29 UK (NI) Nephrops and whitefish vessels and four UK (E\&W) vessels registered in Irish Sea ports at the end of 2001 . Of these, 13 were vessels that used demersal trawls with mesh size $>=100 \mathrm{~mm}$.

## A.3. Ecosystem aspects

Recruitment in Irish Sea whiting appears less variable than in cod and haddock, although there is some similarity in the timing of strong and weak year classes that may indicate a similar response to changes in environmental conditions affecting spawning or earlystage survival. The diet of Irish Sea whiting has been examined in some detail since the 1970s using samples collected from research vessels. Cannibalism occurs in adult whiting; however the effect of this on the assessment of the stock has not yet been investigated. Young whiting are common in the diets of larger predators such as cod and anglerfish.

## B. Data

## B.1. Commercial catch

## B.1.1. Landings

The following table gives the source of landings data for Irish Sea whiting:

|  | Kind of data |  |  |  |  |
| :--- | :---: | :--- | :--- | :--- | :--- |
| Country | Caton (catch- <br> in-weight) | Canum (catch- <br> at-age in <br> numbers) | Weca <br> (weight-at- <br> age in the <br> catch) | Matprop <br> (proportion <br> mature-by-age) | Length <br> composition-in- <br> catch |
| UK(NI) | X | X | X | X | X |
| UK(E\&W) | X | X | X |  | X |
| UK(Scotland) | X |  | X |  | X |
| UK (IOM) | X | X | X |  |  |
| Ireland | X | X |  |  |  |
| France | X |  |  |  |  |
| Belgium | X |  |  |  |  |
| Netherlands | X |  |  |  |  |

Quarterly landings and length/age composition data are supplied from databases maintained by national Government Departments and research agencies. These figures may be adjusted by national scientists to correct for known or estimated misreporting by area or species. Data are supplied on paper or Excel files to a stock coordinator nominated by the ICES Northern Shelf Demersal Working Group, who compiles the international landings and catch-at-age data, and maintains a time-series of such data with any amendments. To avoid double counting of landings data, each UK region supplies data for UK landings into its regional ports, and landings by its fleet into non-UK ports.

The UK (E\&W) currently supplies raised quarterly length frequencies of landings but only sporadic age data. The catch and mean weight-at-age are estimated using combined UK (NI) and Irish quarterly length-weight relationships and age-length keys. Quarterly landings are provided by the UK (Scotland), Belgium and France and annual landings are provided by UK (IOM). The quarterly estimates of landings-at-age into UK (E\&W), UK (NI) and Ireland are raised to include landings by France, Belgium, UK (Scotland), UK (IOM) (distributed proportionately over quarters), and then summed over quarters to produce the annual landings-at-age.

The Excel spreadsheet files used for age distribution, adjustments and aggregations can be found with the stock co-ordinator and for the current and previous year in the ICES computer system under w:\acfm\wgnsds\year\personal\name (of stock co-ordinator).

The result files (FAD data) can be found at ICES and with the stock co-ordinator, as ASCII files on the Lowestoft format, under w:lacfm\wgnsds\yearldatalwhg_7a.

## B.1.2. Discards

The Irish Sea Nephrops fishery takes place on the whiting nursery grounds of the north western Irish Sea and has traditionally produced high whiting discarding. The quantity
of whiting discarded from the UK (NI) Nephrops fishery in 2002 was estimated on a quarterly basis from samples of discards and total catch provided by skippers. The discards samples contain the heads of Nephrops tailed at sea. Using a length-weight relationship, the live weight of Nephrops that would have been landed as tails only is calculated from the carapace lengths of the discarded heads. The number of whiting in the discard samples is summed over all samples in a quarter and expressed as a ratio of the summed live weight of Nephrops in the discard samples (i.e. those represented as heads only in the samples). The reported live weight of Nephrops landed as tails only is then used to estimate the quantity of whiting discarded using the whiting:Nephrops ratio in the discard samples. The length frequency of whiting in the discard samples is then raised to the fleet estimate, and numbers and mean weight-at-age of discarded whiting is computed from the age-length key and length-weight parameters for whiting. The UK (NI) estimates are available since 1980 but the reliability of these estimates has not been determined. Roughly 40 discard samples are collected annually.

There are several limitations to these data: only a small subset of single-rig trawlers is sampled; the method of raising to the fleet discards will be affected by any inaccuracies in the reported landings of Nephrops; and there are no estimates of landings of whiting from these vessels with which to calculate proportions discarded-at-age. However, the WG has used these data in past assessments because removal of discards data would remove a large fraction of catch from the assessment.

A re-analysis of the Irish discard data raised to the Nephrops landings produced estimates of discards from the Irish Nephrops fleet that were more consistent with those of the UK (NI) Nephrops fleet. However, this method of raising could not be used to recalculate an entire time-series of discard estimates from the Irish Nephrops fleet. The quarterly UK (NI) discard ratios were therefore used by the Working Group to estimate the tonnage discarded from the Irish Nephrops fishery. Length frequencies and age-length keys from the whiting discarded by the Irish Nephrops fleet are used to estimate the numbers discarded-at-age from the Irish Nephrops fleet.

At the WGNSDS 2006 revised Irish discard estimates (1996-2005) raised according to the methods described in Borges et al., 2005 were available to the Working Group See Table 1.0. These are available in the ICES files. Discard rates in this series were variable compared with previous estimates based on the UK NI self sampling scheme. Given the differences in raising procedure applied to the NI Discard estimates and the Irish discard estimates further examination of the discard data is needed before international estimates of discard numbers-at-age can be made. The Working Group did therefore not estimate international discard volumes and numbers-at-age for 2004.

## B.2. Biological

Natural mortality was assumed to be constant ( $\mathrm{M}=0.2$, applied annually) for the whole range of ages and years.

A combined sex maturity is assumed, knife-edged at age 2 . The use of a knife-edged maturity ogive has been a source of criticism in previous assessments. However, recent research on gadoid maturity conducted by the UK (NI) gives no evidence for substantial change in whiting maturity since the 1950s, although there has been an increase in the incidence of precocious maturity at age 1, particularly in males, since 1998.

As in previous years, SSB is computed at the start of each year, and the proportions of M and F before spawning were set to zero.

Stock weights are calculated using a procedure first described in the 1998 Working Group Report. To derive representative stock weights for the start of the year for year $i$ and age $j$ the following formula is adopted:

$$
\left(\mathrm{CW}_{i, j}+\mathrm{CW}_{i+1, j+1}\right) / 2=\mathrm{SW} \text { at start of year. }
$$

These values are then smoothed using a 3-year moving average.
Recent investigations into the biological parameters (maturity, sex and growth parameters) of whiting in VIIa (funded under the Data Directive Regulation (1639/2001)) took place during a Biological Sampling survey (BBS) in March 2004. Parameter estimates of maturity at length indicate the L50 for whiting in VIIa for males and females is 13.65 cm and 19.76 cm , respectively. Maturity-at-age for both sexes are similar for most stock area (VIIa, b, j and g) with the notable exception of age 1 males in the Celtic Sea where the estimates are outside the $95 \%$ CI bounds for VIIa and considerably lower than VIa. In most areas whiting were mature by age three and most were mature at age 2 . The sex ratio for whiting tended to increase with length for nearly all the age classes in all areas indicating that females tend to have larger length-at-age than males (Gerritsen, 2005).

Gerritsen et al., 2002 describes the relationships between maturity, length and age of whiting sampled on a length-stratified basis from NI groundfish surveys of the Irish Sea during spawning in spring 1992-2001. Findings show that most one year old females were immature whilst most two year old females were mature; almost all 3 year olds of both sexes were mature. Length at 50 maturity average around 19 cm in males and 22 cm in females.

## B.3. Surveys

Seven research vessel survey-series for whiting in VIIa were available to the Working Group in 2005. In all surveys listed the highest age represents a true age not a plus group.

- UK (England and Wales) Beam Trawl Survey (UK E\&W-BTS): ages 0 and 1, years 1988-2002: The survey covers the entire Irish Sea and is conducted in September on the R.V. Corystes. The survey uses a 4 m beam trawl targeted at flatfish. The survey is stratified by area and depth band, although the survey indices are calculated from the total survey catch without accounting for stratification. Numbers of whiting at age per km towed are provided for prime stations only (i.e. those fished in most surveys).
- UK (Northern Ireland) October Groundfish Survey (NIGFS-October): ages 0-5, years 1992-2005: The survey series commenced in its present form in 1992. It comprises 45 three mile tows at fixed station positions in the northern Irish Sea, with an additional 12 one mile tows at fixed station positions in the St George's channel from October 2001 (the latter are not included in the tuning data). The surveys are carried out using a rockhopper otter trawl deployed from the R.V. Lough Foyle. The survey designs are stratified by depth and sea bed type. The mean numbers at length per three mile tow are calculated separately by stratum, and weighted by surface area of the strata to give a weighted mean for the survey or group of strata. The strata are grouped into
western Irish Sea and eastern Irish Sea, and a separate age-length key is derived for each area to calculate abundance indices by age class. The survey design and time-series of results including distribution patterns of whiting are described in detail in Armstrong et al., 2003.
- UK (Northern Ireland) March Groundfish Survey (NIGFS-March): ages 1-5, years 1992-2006: Description as for UKNI-GFS-October above.
- UK (Northern Ireland) Methot Isaacs-Kidd Survey (UKNI-MIK): age 0, years 1993-2005: The survey uses a Methot Isaacs-Kidd frame trawl to target pelagic juvenile gadoids in the western Irish Sea at $40-45$ stations. The survey is stratified and takes place in June during the period prior to settlement of gadoid juveniles. Indices are calculated as the arithmetic mean of the numbers-per-unit sea area.
- Ireland's Irish Sea Celtic Sea Groundfish Survey (IR-ISCSGFS): ages $0-5$, years 1997-2002: This survey commenced in 1997 and is conducted in OctoberNovember on the R.V. Celtic Voyager. The $\alpha$ and $\beta$ of the series are set to account for the variable timing of this survey within the fourth quarter. The survey uses a GOV otter trawl with standard ground gear and a 20 mm codend liner. The survey operates mainly in the western Irish Sea but has included some stations in the eastern Irish Sea. The survey design has evolved over time and has different spatial coverage in different years. Indices are calculated as arithmetic means of all stations, without stratification by area.
- UK (Scotland) groundfish survey in spring (ScoGFS-spring): ages 1-8, years 1996-2006: This survey represents an extension of the Scottish West Coast groundfish survey (Area VI), using the research vessel Scotia. The survey gear is a GOV trawl, and the design is two fixed-position stations per ICES rectangle from 1997 onwards ( 17 stations) and one station per rectangle in 1996 (nine stations). The survey extends from the Northern limit of the Irish Sea to around $53^{\circ} 30^{\prime}$.
- UK (Scotland) groundfish survey in Autumn (ScoGFS-autumn): ages 0-5, years 1997-2005: The survey covers a similar area to the ScoGFS in spring, but has only 11-12 stations.
- IRGFS (Ireland): This survey commenced in 2003 aboard the R.V. Celtic Explorer. It is a depth stratified survey using a GOV trawl with a 20 mm mesh liner on the codend. The survey currently covers VIIb, j, g and VIa. Prototcols for the survey are governed by the International Bottom Trawl Survey Working Group (IBTS).

To allow the inclusion of the NIGFS-March and ScoGFS-spring surveys for the year after the last year with commercial catch data in an XSA, the surveys may be treated as if they took place at the end of the previous year, and the age range and year range of the surveys may be shifted back accordingly in the data files.

The following research surveys were available to the 2007 Working Group:

- UK (NI) groundfish survey: March 1992-2007.
- UK (NI) groundfish survey: October 1992-2006.
- UK (Scotland) groundfish survey: March 1996-2006.
- UK (Scotland) groundfish survey: autumn 1997-2005.
- Irish groundfish survey: autumn 2003 and 2004.
- UK (NI) MIK net surveys of pelagic-stage 0-group cod, western Irish Sea 19942006.
- UK (E\&W) beam trawl survey: 0-1 gp cod, 1988-2006.

FSP surveys of Irish Sea round fish: 2004-2007.
Further details of the tuning data are given in Appendix 1 and 2 of the 1999 WG Report.

## B.4. Commercial cpue

No cpue data have been provided for the French (Lorient) trawl fleet since 1992. Four commercial catch-effort dataseries were available to the WG:

- Irish otter trawl (IR-OTB): ages 1-6, years 1995-2002: Effort and cpue data provided for the Irish fleet comprise total annual effort (hours fished, not corrected for fishing power) and total numbers-at-age in landings from otter trawlers. The data were revised to take account of updated logbook information. This fleet operates mainly in the western Irish Sea, targeting Nephrops and/or whitefish. The distribution of fishing is concentrated in the western part of the range of the whiting stock in the Irish Sea. Hence the catch rates will represent changes in abundance of whiting in the western part of VIIa. The use of this fleet as a tuning index therefore relies on the assumption that trends in abundance in the west of VIIa reflect those of the entire stock. The catch-at-age data comprise a large proportion of the total international catch. Hence, some correlation of errors can be expected between the tuning dataset and the catch-at-age data. The effect of such correlations has not been evaluated. The otter trawl catch-at-age data contained data for landings only. Hence the reliability of the tuning fleet will be limited for age groups which are heavily discarded.
- UK (Northern Ireland) pelagic trawl: ages 2-6, years 1993-2002: The pelagic trawl catch-at-age data contained data for landings only. Hence the reliability of the tuning fleet will be limited for age groups which are heavily discarded. This fleet currently targets haddock and cod in the deeper waters of the western Irish Sea and the North Channel. Bycatches of whiting are currently very small and are heavily discarded due to their low value. The fleet is considered unsuitable for indexing whiting abundance.
- UK (Northern Ireland) single rig otter trawl: ages 0-6, years 1993-2002: This fleet operates mainly in the western Irish Sea. The distribution of fishing does not encompass the entire range of the whiting stock (which surveys suggest is distributed across the Irish Sea). Whiting discards from single-rig trawlers (estimated from fisher self-sampling scheme) are included.
- UK (England and Wales) otter trawl: ages 2-6, years 1981-2000: Estimates up to and including 2000 of commercial lpue from UK (E\&W) otter trawlers contain data for landings only. Hence the reliability of the tuning fleet will be limited for age groups which are heavily discarded. This fleet operates mainly in the eastern Irish Sea. The distribution of fishing does not encompass the entire
range of the whiting stock (which surveys suggest is distributed across the Irish Sea) or the main whiting nursery grounds (in the western Irish Sea). Age compositions in most years have been estimated from length frequencies using ALKs that were obtained from sampling of fleets operating mainly in the western Irish Sea. This has introduced additional uncertainties into the data.


## B.5. Other relevant data

None.

## C. Historical stock development

Model used:
XSA (up to 2002)
SURBA 2.0-2003
SURBA 3.0-2004
SURBA 2.2-2005
Software used:
Lowestoft VPA suite
XSA Model Options chosen:
Tapered time weighting not applied
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=4$
Survivor estimates shrunk towards the mean F of the final 5 years or the 2 oldest ages
S.E. of the mean to which the estimate are shrunk $=0.500$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied

Input data types and characteristics:

|  | Name |  |  | Variable from <br> year to year |
| :--- | :--- | :--- | :---: | :--- |
| Type | Catch in tonnes | 1980-last data <br> year | $0-6+$ | Yes |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :--- | :---: |
| Tuning fleet 1 | NIGFS-Oct | 1992-last data year | $0-5$ |
| Tuning fleet 2 | NIGFS-Mar (adjusted) | 1991-(last data year-1) | $0-4$ |
| Tuning fleet 3 | ScoGFS-Spring | 1996-last data year | $1-5$ |
| Tuning fleet 4 | UK(E\&W) BTS | 1988-last data year | $0-1$ |

For analysis of alternative procedures see WG reports from WGNSDS 1997-2005.

## D. Short-term projection

Model used:
Age structured
Software used:
MFDP prediction with management option table and yield-per-recruit routines. MLA suite (WGFRANSW) used for sensitivity analysis and probability profiles.

Initial stock size.

Taken from the XSA for age 1 and older. The recruitment at age 0 in the last data year is estimated as a short-term GM (1992 onwards) because of a reduction in mean recruitment since then.

Natural mortality:
Set to 0.2 for all ages in all years.
Maturity:
The same ogive as in the assessment is used for all years.
F and M before spawning:
Set to 0 for all ages in all years.
Weight-at-age in the stock:
average stock weights for last three years.
Weight-at-age in the catch:
Average weight of the three last years.

## Exploitation pattern:

Average of the three last years. Discard F's, which are generated by the Nephrops fleet as there are no discard estimates for other fleets, are held constant while landings $\mathrm{F}^{\prime}$ s are varied in the management option table.

Intermediate year assumptions:
status quo F
Stock-recruitment model used:
None, the short-term geometric mean recruitment at age 0 is used.
Procedures used for splitting projected catches:
F vectors in each of the last three years of the assessment are multiplied by the proportion landed or discarded-at-age to give partial Fs for landings and discards. The vectors of partial Fs are then averaged over the last three years to give the forecast values.

## E. Medium-term projections

No medium-term projections are done for this stock due to problems with estimating current F .

## F. Yield and biomass-per-recruit/long-term projections

Model used:
yield and biomass-per-recruit over a range of F values that may reflect fixed or variable discard F's.

Software used:

MFY or MLA
Selectivity pattern:
mean F array from last 3 years of assessment (to reflect recent selection patterns).
Stock and catch weights-at-age:
mean of last three years (weights-at-age have declined as the stock has declined since the 1980s; it is not known if this is an environmental effect on growth that is independent of stock size).

Proportion discarded:
Partial F vectors are the recent average.
Maturity:
Fixed maturity ogive as used in assessment.

## G. Biological reference points

Precautionary approach reference points have remained unchanged since 1999. $\mathrm{B}_{\mathrm{pa}}$ is set at 7000 t and is defined as $\mathrm{B}_{\lim }{ }^{*} 1.4$. Blim is defined as the lowest observed SSB (ACFM, 1999), considered to be 5000 t . There is not considered to be clear evidence of reduced recruitment at the lowest observed SSBs. $\mathrm{F}_{\mathrm{pa}}$ is set at 0.65 on the technical basis of high probabilities of avoiding $\mathrm{F}_{\mathrm{lim}}$ and of SSB remaining above $\mathrm{B}_{\mathrm{pa}}$ in the long term. Flim is defined as 0.95 , the fishing mortality estimated to lead to a potential stock collapse.

## H. Other issues

None.

## I. References

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Borges, L.; Rogan, E. and Officer, R. 2005. "Discarding by the demersal fishery in the waters around Ireland", Fish. Res. (in press).

Gerritsen, H. 2005. Biological parameters for Irish Demersal Stocks in 2004. WD5 (WGNSDS, 2005)

Table 1.0 Revised Discard estimates raisesd according to the method oulined in Borges et al., 2005.

\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline \& \multirow[t]{2}{*}{\begin{tabular}{l}
1996
Numbers Weight ('000) \\
(kg)
\end{tabular}} \& \multirow[t]{2}{*}{\begin{tabular}{|c|c|}
\hline Numbers Weigh \\
('000) \& (kg)
\end{tabular}} \& \multirow[t]{2}{*}{} \& \multirow[t]{2}{*}{\begin{tabular}{|c|}
\hline \multicolumn{2}{|c|}{1999} \\
NumbersWeigh \\
\((' 000)\) \\
\((\mathrm{kg})\)
\end{tabular}} \& \multicolumn{2}{|l|}{2000} \& \multicolumn{2}{|l|}{2001} \& \multicolumn{2}{|l|}{2002} \& \multicolumn{2}{|l|}{2003} \& \multicolumn{2}{|l|}{2004} \& \multicolumn{2}{|l|}{2005} \\
\hline Age \& \& \& \& \& Numbers ('000) \& \[
\begin{gathered}
\text { Weight } \\
(\mathrm{kg}) \\
\hline
\end{gathered}
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\end{gathered}
\] \& Numbers ('000) \& \[
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\text { Weight } \\
(\mathrm{kg})
\end{gathered}
\] \\
\hline 0 \& \(5631.20 \quad 0.015\) \& 4110.630 .027 \& 5073.570 .027 \& 187.260 .036 \& 7850.12 \& 0.033 \& 20981.54 \& 0.016 \& 29017.16 \& 0.021 \& 1921.76 \& 0.016 \& 17091.56 \& 0.018 \& 442.07 \& 0.010 \\
\hline 1 \& \(5925.33 \quad 0.035\) \& 8361.190 .044 \& 5939.530 .064 \& 276.500 .102 \& 3098.24 \& 0.047 \& 8883.11 \& 0.054 \& 12097.93 \& 0.033 \& 2419.56 \& 0.036 \& 7347.29 \& 0.034 \& 2531.84 \& 0.035 \\
\hline 2 \& 1802.900 .111 \& 3243.450 .120 \& 3826.200 .107 \& 150.990 .174 \& 137.80 \& 0.153 \& 1413.48 \& 0.126 \& 576.17 \& 0.112 \& 1287.21 \& 0.178 \& 731.35 \& 0.101 \& 783.68 \& 0.091 \\
\hline 3 \& \(144.34 \quad 0.217\) \& 696.180 .200 \& 440.050 .185 \& \(43.70 \quad 0.235\) \& 30.31 \& 0.229 \& 479.38 \& 0.133 \& 152.95 \& 0.105 \& 603.20 \& 0.246 \& 142.50 \& 0.165 \& 129.28 \& 0.159 \\
\hline 4 \& \(6.02 \quad 0.206\) \& 68.710 .241 \& 0.000 .000 \& 0.000 .000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 108.64 \& 0.268 \& 96.30 \& 0.218 \& 40.12 \& 0.154 \\
\hline 5 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.00 \& 0.000 \& 22.95 \& 0.136 \& 17.66 \& 0.123 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 24.48 \& 0.371 \\
\hline 6 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.000 .000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \\
\hline 7 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \\
\hline 8 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.000 .000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \\
\hline 9 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \\
\hline 10 \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& \(0.00 \quad 0.000\) \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \& 0.00 \& 0.000 \\
\hline \multirow[t]{2}{*}{OTB Discards (tonnes, whole weight)} \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \& \\
\hline \& 520.8 \& \#\#\# \& 1010.3 \& 71.6 \& \multicolumn{2}{|r|}{434.3} \& \multicolumn{2}{|r|}{1054.5} \& \multicolumn{2}{|r|}{1100.9} \& \multicolumn{2}{|r|}{523.6} \& \multicolumn{2}{|r|}{680.3} \& \multicolumn{2}{|r|}{201.3} \\
\hline \multirow[t]{2}{*}{Sampling Information Number of Trips} \& 1996 \& 1997 \& 1998 \& \multirow[t]{3}{*}{1999 4} \& \multicolumn{2}{|r|}{\multirow[t]{3}{*}{\(2000 \begin{array}{rr} \\ \& 10 \\ \& 111\end{array}\)}} \& \multicolumn{2}{|r|}{\multirow[t]{3}{*}{\(2001 \begin{array}{rr} \\ \& 2 \\ \& 34\end{array}\)}} \& \multicolumn{2}{|l|}{\multirow[t]{3}{*}{2002}} \& \multicolumn{2}{|r|}{\multirow[t]{3}{*}{2003

60}} \& \multicolumn{2}{|r|}{\multirow[t]{3}{*}{$2004 \begin{array}{rr} \\ & 11 \\ & 122\end{array}$}} \& \multicolumn{2}{|r|}{\multirow[b]{3}{*}{8
96}} <br>
\hline \& \multirow[t]{2}{*}{8

48} \& $$
8
$$ \& \& \& \& \& \& \& \& \& \& \& \& \& \& <br>

\hline Number of Hauls \& \& 44 \& 58 \& \& \& \& \& \& \& \& \& \& \& \& \& <br>
\hline
\end{tabular}

## Stock Annex 6.7: Irish Sea Plaice

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Plaice (Division VIIa) |
| :--- | :--- |
| Working Group | Celtic Seas Ecoregion |
| Date | 18th May 2010 |
| By | Christopher Lynam |

## A. General

## A.1. Stock definition

There are considered to be three principle spawning areas of plaice in the Irish Sea: one off the Irish coast, another northeast of the Isle of Man towards the Cumbrian coast, and the third off the north Wales coast (Nichols et al., 1993; Fox et al., 1997; Figure A1). Cardigan Bay has also been identified as a spawning ground for plaice in the Irish Sea (Simpson, 1959).

The level of mixing between the east and west components of the Irish Sea stock appears small. (Dunn and Pawson, 2002). Length-at-age measurements from research surveys as well as anecdotal information from the fishing industry suggests that plaice in the western Irish Sea grow at a much slower rate than those in the eastern Irish Sea. Earlier studies have suggested that the east and west components of the stock are distinct (Brander 1975; Sideek 1989) and should therefore be considered independently of one another. Morphometric differences have been observed between the east and west components of the stock; a comment in the 1982 WG report states that plaice to the west of the $5^{\circ} \mathrm{W}$ line are approximately 3 cm larger at-age (for the most abundant age groups) than those to the east of this line. In contrast, the 2004 WG indicated that the UK (E\&W) beam trawl survey in September (from 1989) catches plaice off the Irish coast that are smaller-at-age than those caught in the eastern Irish Sea. In 2009, the raw catch weight data (prior to polynomial smoothing) from UK (E\&W) and Irish fleets (all gears) indicates that plaice caught by the Irish fleets are approximately 50 g heavier than those caught by the UK (E\&W) fleet (Figure A2).

The degree of separation between the stocks of plaice in the Irish Sea and the Celtic Sea is unclear. Numerous tagging studies indicate a southerly movement of mature fish (or fish maturing for the first time) from the southeast Irish Sea, off North Wales, into the Bristol Channel and Celtic Sea during the spawning season, such that $43 \%$ of the new recruits are likely to recruit outside of the Irish Sea (Figure A1). While some of these migrant spawning fish will remain in the Bristol Channel and Celtic Sea, the majority ( $\ell 70 \%$ ) are expected to return to summer feeding grounds in the Irish Sea (Dunn and Pawson, 2002).

Very little mixing is considered to occur between the Irish Sea and Channel stocks or between the Irish Sea and North Sea (Pawson 1995). Nevertheless, time-series of recruitment estimates for all stocks in waters around the UK (Irish Sea, Celtic Sea, western and eastern Channel, North Sea) show a significant level of synchrony (Fox et al., 2000). This could indicate that the stocks are subject to similar large-scale environmental forces and respond similarly to them, or alternatively that there are subpopulations that share a common spawning.


Figure A1. (right) Principal substock areas and movements of plaice on the west coast of England and Wales. Percentages are the recaptures rates of tagged plaice $<25 \mathrm{~cm}$ total length when released, and $>26 \mathrm{~cm}$ when recaptured in English and Welsh commercial fisheries. Tagging exercises in 1979-1980 and 1993-1996 were combined based on the assumption that the dispersal patterns of plaice were consistent over time. For each substock, the main feeding area (derived from tag recaptures during April-December; light shading), and the main spawning area (derived from tag recaptures during January-March, and ichthyoplankton surveys; dark shading) are indicated. The substocks tagged have been coloured green, red and blue. The substocks coloured orange are less well determined, with the feeding area around south-east Ireland unknown. Letters represent return migrations, where $A \approx 6 \%$, and $B+C \approx 46 \%$. Reproduced from Dunn and Pawson (2002).


Figure A2. Observed weight-at-age of plaice from landed catches by the UK (E\&W) and Irish fleets (all gears) in 2009.

## A.2. Fishery

The status and activities of the fishing fleets operating in ICES Subdivision VIIa are described by Pawson et al., 2002 and also by Anon, 2002. The majority of vessels operating in the Irish Sea are otter trawlers fishing for cod, haddock, whiting and plaice with bycatch of anglerfish, hake and sole. Since 2001 these trawlers have adopted mesh sizes of $100-120 \mathrm{~mm}$ and other gear modifications depending on the requirements of recent EU technical conservation regulations and national legislation. Square mesh panels have been mandatory for UK otter trawlers since 1993 and for Irish trawlers since 1994. The number of Irish vessels operating in this area has declined in recent years. Fishing effort (hours fished) in the UK (England and Wales) fleet declined rapidly after 1989 and in 2009 effort by the Irish and UK (E\&W) otter fleets reached historic lows.

Although some of the otter trawlers also take part in the fishery for sole, there have been a growing number of beam trawlers, particularly from southern England and Belgium exploiting sole. This fishery has important bycatch of plaice, rays, brill, turbot and anglerfish. The fishing effort of the Belgium beam trawl fleet varies according to the catch rates of sole in the Irish Sea compared with other areas in which the fleet operates. In 2009, effort (hours fished) by the UK (E\&W) beam trawl fleet fell to the lowest observed level.

A fleet of vessels primarily from Ireland and Northern Ireland take part in a targeted Nephrops fishery using 70 mm mesh nets with 75 mm square mesh panels. This fishery takes a substantial bycatch of whiting, most of which is discarded. Some inshore shrimp beam trawlers occasionally switch to flatfish when shrimp become temporarily unavailable. Other gear types employed in the Irish Sea to catch demersal species are gillnets and tanglenets, notably by inshore boats targeting cod, bass, grey mullet, sole and plaice.

The minimum landing size for plaice in the Irish Sea was set in 1980 to 25 cm (Council Regulation (EEC) No 2527/80). This was increased in 1998 to 27 cm (Annex XII of Council Regulation 850/98).

Since 2000 a recovery programme has been implemented to reduce exploitation of the cod spawning stock in the Irish Sea. In 2002 the European Commission regulations included a prohibition on the use of demersal trawl, enmeshing nets or lines within the main cod spawning area in the northwest Irish Sea between the 14th February
and 30th April. Some derogations were permitted for Nephrops trawls and beam trawlers targeting flatfish.

## A.3. Ecosystem aspects

Plaice are preyed upon and consume a variety of species through their life history. However, plaice have not as yet been included in an interactive role in multispecies assessment methods (e.g. ICES WGSAM 2008). Among other prey items, plaice typically consume high proportions of polychaetes and molluscs.

Other than statistical correlations between recruitment and temperature (Fox et al., 2000) little is known about the effects of the environment on the stock dynamics of plaice in the Irish Sea. Negative correlations between year-class strength of plaice (in either the Irish Sea, Celtic Sea, Channel and North Sea) and sea surface temperature are generally strongest for the period February-June. However, western (North Sea and Channel) and eastern (Irish Sea and Celtic Sea) stocks have been found to respond to different time-scales of temperature variability, which might imply that different mechanisms are operating in these stocks and/or that the Irish Sea and Celtic Sea share common spawning (Fox et al., 2000).

## B. Data

## B.1. Commercial catch

## Landings

International landings-at-age data based on quarterly market sampling and annual landings figures are available from 1964. Throughout the period 1978 to 2003 quarterly age compositions have typically represented around $80-90 \%$ of the total international landings. Table B1 details the derivation of international landings for the period 1978 to 2003.

Prior to 1983 the stock was assessed on a separate sex basis: the catch numbers of males and females were worked up separately and the numbers of males and females in the stock as estimated from each assessment combined to give a total biomass estimate. Since 1983 a combined sex assessment of the stock has been conducted and the numbers of males and females in the catch have been combined at the international data aggregation level prior to running a single assessment.

## Data exploration

Data exploration for commercial landings data for Irish Sea plaice currently involves:

- expressing the total landings-at-age matrix as proportions-at-age, normalised over time, so that year classes making above-average contributions to the landings are shown as large positive residuals (and vice-versa for be-low-average contributions);
- applying a separable VPA model in order to examine the structure of the landed numbers-at-age before they are used in catch-at-age analyses, in particular whether there are large and irregular residuals patterns that would lead to concerns about the way the recorded catch has been processed.


## Discards

In 1986, the UK fleet was restricted to a $10 \%$ bycatch of plaice for almost the entire year. Estimates were made of the increased quantity of plaice that would have been discarded based on comparisons of lpue values for 1985-1986 with those for 19841985. The estimated quantity of 250 tonnes was added to the catch. A similar situation arose the following year and 250 tonnes was added to the catch for 1987.

The $10 \%$ plaice bycatch restriction was enforced again in 1988 to all UK (E\&W) vessels in the 1st quarter and to beam trawlers in the 2nd and 3rd quarters. However, this time the landings were not corrected for discard estimates.

Discard information is not routinely incorporated into the assessment.

## B.2. Biological

## Weights-at-age

A number of different methodologies have been employed to determine weights-atage for this stock. Stock weights and catch weights-at-age were determined on a separate sex basis and remained unchanged from 1978 until 1983. Catch weights were derived from a von Bertalanffy length-at-age fit to Belgian (70-74), UK (E\&W) (64-74) and Irish (62-66) catch samples. The estimated lengths-at-age were converted to weights-at-age using a Belgian length-weight dataset (ages 2-15 females; 3-9 males). Stock weights were calculated as the mean of adjacent ages from the catch weights, where catch weights represented 1st July values and stock weights 1st January.

From 1983 weights-at-age have been calculated on a combined sex basis. Catch weights were taken from market sampling measurements combined on a sex weighted basis and smoothed. For the period 1983 to 1987 catch weights were smoothed by eye, from 1988 onwards a smooth curve was fitted using a numerical minimization routine. Stock weights were derived from the smoothed international catch weights-at-age curve with values representing 1st January. In 1985 the stock weights-at-age were adjusted for ages 1 to 4 . The difference between the smoothed catch weights and survey (F.V. Silver Star) observations were adjusted using the maturity ogive to give "best estimate" stock weights "for ages where growth and maturity differences can bias sampling procedures". The same procedure was adopted in 1986 (when stock weights in 1982 and 1983 were also revised so as to be consistent with this methodology) and 1987. In 1988 however, the Silver Star survey was discontinued and stock weights-at-ages 1 to 3 were calculated as means of the three previous years. Correction of the estimated stock weights of the younger age groups did not occur in 1989 or in subsequent years which explains the sudden increase in weight of the younger age groups for this stock from 1988 onwards.

Catch weights at the younger ages also show a similar increase coincident with the start of the smoothing process. This apparent increase in the estimated catch weights is not believed to have affected the derivation of catch numbers since smoothing of the catch weights occurs after having determined the catch numbers-at-age. SOP checks are generally very close to $100 \%$.

The 1982 WG Report notes a study by R. Cross, unpublished stating that there was no evidence for a change in growth rates for the stock nor was there any evidence of density-dependent effects on growth.

## Natural mortality and maturity ogives

As for the weights-at-age, natural mortality and maturity was initially determined on a separate sex basis. Natural mortality was taken as 0.15 for males and 0.1 for females. In 1983 when a combined sex assessment was undertaken a sex weighted average value of 0.12 was used as an estimate of natural mortality. This estimate of natural mortality has remained unchanged since 1983.

The maturity estimates used prior to 1982 are not specified. A new separate sex maturity ogive (Sideek, 1981) was implemented in 1982. This ogive was recalculated as sex weighted mean values in 1983 when the assessment was conducted on a combined sex basis. The maturity ogive was revised again in 1992 based on the results of an EU project. Maturity ogives are applied as vectors to all years in the assessment.

Table A1. Maturity ogives for Irish Sea plaice used in ICES WGs.

| Age | WG 1978-1982 |  | WG 1983-1992 | WG 1992-2009 |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | M | F |  |  |
| 1 | 0 | 0 | 0 | 0 |
| 2 | 0.3 | 0.04 | 0.15 | 0.24 |
| 3 | 0.8 | 0.4 | 0.53 | 0.57 |
| 4 | 1.0 | 0.94 | 0.96 | 0.74 |
| 5 | 1.0 | 1.0 | 1.0 | 0.93 |
| 6 | 1.0 | 1.0 | 1.0 | 1.0 |

The proportion of fishing mortality and natural mortality before spawning was originally set to 0 . It was changed in 1983 to a value of 0.2 on the grounds that approximately $20 \%$ of the catch was taken prior to March (considered to be the time of peak spawning activity). As for Celtic Sea plaice the proportion of F and M before spawning was reset to 0 , as it was considered that these settings were more robust to changes in the fishing pattern, especially with respect to the medium-term projections.

## B.3. Surveys

In 1993, the UK (E\&W) beam trawl survey-series that began in 1988 was considered to be of sufficient length for inclusion in the assessment. Since 1991, tow duration has been 30 minutes but prior to this it was 15 minutes. In 1997, values for 1988 to 1990 were raised to 30 minute tows. However, data for 1988 and 1989 were of poor quality and gave spurious results: thus, the series was truncated to 1990. A similar March beam trawl survey began in 1993 and was made available to the WG in 1998. The March beam trawl survey ended in 1999 but continued to be used as a tuning index in the assessment until 2003.

An Irish juvenile plaice survey index was presented to the WG in 2002 (1976-2001, ages 2-8). Between 1976 and 1990 this survey had used an average ALK for that period. Serious concerns were expressed regarding the quality of the data for this period and the series was truncated to 1991. The stations for this survey are located along the coast of southeast Ireland between Dundalk Bay and Carnsore Point and there was some concern that this localised survey-series would not be representative of the plaice population over the whole of the Irish Sea. Numerous tests were conducted at the 2002 WG to determine the validity of this and other tuning indices and
it was concluded that this survey could be used as an index of the plaice population over the whole of the Irish Sea.

The SSB of plaice can be estimated using the Annual Egg Production Method (AEPM) (Armstrong et al., 2002 and WD 11, WGCSE 2010). This method uses a series of ichthyoplankton surveys to quantify the spatial extent and seasonal pattern of egg production, from which the total annual egg production can be derived. The average fecundity (number of eggs spawned per unit body weight) of mature fish is estimated by sampling adult females immediately prior to the spawning season. Dividing the annual egg production by average fecundity gives an estimate of the biomass of mature females. Total SSB can be estimated if the sex ratio is known. Although substantial discrepancies between absolute estimates of SSB from the Annual Egg Production method (AEPM) and the ICES catch-based assessments were observed, they do confirm that SSB of plaice in the Irish Sea is currently at high levels.

AEPM estimates of SSB for plaice (RSE = relative standard error, as \%), based on production of Stage 1 eggs) are shown below (note 1995-2006 estimates were revised in 2010, see WD11 WGCSE 2010):

Table A3. AEPM estimates of SSB for Irish Sea plaice.

|  | total | west |  | east |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | SSB $(\mathrm{t})$ | RSE | SSB(t) | RSE | SSB( t$)$ | RSE |
| 1995 | 9081 | 21 | 3411 | 42 | 5670 | 22 |
| 2000 | 13303 | 19 | 5654 | 36 | 7649 | 19 |
| 2006 | 11487 | 16 | 3655 | 29 | 7833 | 19 |
| 2008 | 12729 | 19 | 4309 | 43 | 8420 | 18 |

Splitting the SSB estimate by substrata (Figure below) suggests that the perceived increase in plaice SSB is limited to the eastern Irish Sea. This finding agrees with an analysis of UK (NI) GFS data by substrata, which also indicates an increase in biomass limited to the eastern Irish Sea.


Figure A3. AEPM estimates by year and substrata.

## B.4. Commercial cpue

Prior to 1981 tuning data were not used in the assessment of this stock. A separable assessment method was used and estimates of terminal S and F were derived iteratively based on an understanding of the recent dynamics of the fishery.

In 1981 the choice of terminal F was determined from a regression of exploited stock biomass on cpue. Catch and effort series were available for the UK (E\&W) trawl fleet and the Belgian beam trawl fleet for the period 1964 to 1980. In 1994 the Belgian and UK cpue series were combined to provide one mean standardised international index. The UK (E\&W) trawl series was revised in 1986 (not known how) and in 1987 was recalculated as an age based cpue index enabling the use of the hybrid method of tuning an ad hoc VPA.

The UK (E\&W) trawl tuning-series was revised in 1999 and separate otter trawl and beam trawl tuning-series were produced using length samples from each gear type and an all gears ALK. Since the data could only be separated for 1988 onwards the two new tuning-series were slightly reduced in length. In 1996 UK (E\&W) commercial effort data were re-scaled to thousands of hours so as to avoid numerical problems associated with low cpue values and in 2000 the UK (E\&W) otter trawl series was re-calculated using otter trawl age compositions only rather than combined fleet age compositions as previously.

Two newly revised survey indices for the Lough Beltra were presented to the WG in 1996 though they were considered too noisy for inclusion in the assessment. They were revised again for the following year and found to be much improved but were again not included because they ended in 1996 and the WG felt that they would add little to the assessment. An Irish otter trawl tuning index was made available in 2001 (1995-2000, age 0 to 15). Whilst this fleet mainly targets Nephrops, vessels do on occasion move into areas where plaice are abundant. Landings of plaice by this fleet were approximately $15 \%$ of total international landings in 2000 and the WG considered that this fleet could provide a useful index of abundance for plaice.

The effects of vessel characteristics on lpue for UK (E\&W) commercial tuning-series was investigated in 2001 to investigate the requirement for fishing power corrections due to MAGP IV re-measurement requirements. It was found that vessel characteristics had less effect on lpue than geographic factors and unexplained noise and concluded that corrections were not necessary. However, vessels of certain size tended to fish in certain rectangles. This confounding may have resulted in the underestimation of vessel effects.

Currently, age-based tuning data available for this assessment comprise three commercial fleets; the UK (E\&W) otter trawl fleet (UK (E\&W) OTB, from 1987), the UK (E\&W) beam trawl fleet (UK (E\&W) BT, from 1989) and the Irish otter trawl fleet (IR-OTB, from 1995). However, as a consequence of inconsistencies in these commercial tuning fleets and surveys in the Irish Sea no commercial tuning information is used in the assessment.

## B.5. Other relevant data

## C. Historical stock development

The stock of plaice in the Irish Sea has been assessed by ICES since 1977.

## Assessment methods and settings

In 1987 the stock was assessed using a Laurec-Shepherd (hybrid) tuned VPA. Concerns about deteriorating data quality prompted the use in 1994 of XSA. A subsequent divergence in commercial cpue and survey data, and the wish to include biomass indices, prompted the use of ICA. The settings for each of the assessments between 1991 and 2009 are detailed in Table B.2. Since 2006, the assessment has been an update ICA assessment with the separable period increased by one year at each assessment working group. Since 2009, FLICA has been used to run the assessment: the R and FLR packages have been documented within the WG Report.

Over the years, trial runs have explored many of the options with regards XSA settings, including:

- The applicability of the power model on the younger ages was explored in: 1994; 1996; 1998; 1999; 2000 and 2001.
- Different levels of F shrinkage were explored in 1994; 1995; 1997.
- The effect of different time tapers was investigated in 1996.
- The S.E. threshold on fleets was examined in 1996.
- The level of the catchability plateau was investigated in 1994.

ICA settings explored since 2005 have included:

- The length of the separable period;
- The reference age;
- The age range of the landings data.


## D. Short-term projection

Short-term projections are considered unreliable and although presented in the Report they should not be used for advice.

## Software: Multi Fleet Deterministic Projection (MFDP)

Age-based short-term projections are conducted for a three year period using initial stock numbers derived from ICA analyses. Numbers-at-age 2 are considered poorly estimated and are generally overwritten using a geometric mean (GM) of past recruitment values. Population numbers-at-age 3 in the intermediate year (terminal year +1 ) are also overwritten with the GM estimate depreciated for $\mathrm{F}_{s q}$ and natural mortality. Recent recruitments have been estimated to be at a lower level and to be less variable than those earlier in the time-series. Consequently a short-term geometric mean (from 1990 to 2 years before the terminal year) is used.

Currently, the exploitation pattern is an un-scaled three year arithmetic mean. However, alternative options may be used depending on recent F trajectories and the Working Group's perception of the fishery. Catch and stock weights-at-age are generally taken as the mean of the last three years and the maturity ogive and natural mortality estimates are those used in the assessment method.

## E. Medium-term projections

Medium-term projections are not carried out for this stock.
Previous Software: MLA miscellany

Input values to the medium-term forecast were the same as those used in the shortterm forecast. Although a Beverton-Holt stock-recruit relationship has been assumed previously, a simple geometric mean may now be more appropriate. It remains unclear whether the full time-series or a reduced time-series from 1989 should be used.

## F. Yield and biomass-per-recruit/long-term projections

Software: Multi Fleet Yield-per-Recruit (MFYPR)
Yield-per-recruit calculations are conducted using the same input values as those used for the short-term forecasts. Currently the YPR calculations are used as a basis for determining the catch option for advice.

## G. Biological reference points

Biological reference points were proposed for this stock by the 1998 Working Group as below:

|  | Type | Value | Technical basis |
| :--- | :--- | :--- | :--- |
|  | Blim | Not <br> defined. | There is no biological basis for defining Blim as the stock- <br> recruitment data are uninformative. |
| Precautionary <br> approach | $\mathrm{B}_{\mathrm{pa}}$ | 3100 t | $\mathrm{B}_{\mathrm{pa}}=$ Bloss. |

## Yield and spawning biomass-per-Recruit

$F$-reference points:

|  | Fish Mort | Yield/R | SSB/R |
| :--- | :--- | :--- | :--- |
|  | Ages 3-6 |  |  |
| Average last 3 years | 0.10 | 0.17 | 1.64 |
| Fo.1 $^{\text {Fmed }}$ | 0.14 | 0.19 | 1.31 |

Estimated by the WG in 2010

MSY reference points were explored by WGCSE 2010 using the Cefas ADMB code presented to WKFRAME (ICES, 2010). However, due to the high level of discards in the stock and unreliable estimates of recruitment, MSY reference points were rejected by the Working Group.

## H. Other issues

None.

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Table B.1. Data sources and derivation of international landings, where $\%$ sampled indicates the percentage of the total landings represented by sampling.

| Year |  | Source |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| of WG | Data | UK | Belgium | Ireland | Netherland | Derivation of international landings | $\%$ <br> sampled |
| 1978 | Len. comp. | quarterly1quarterly1quarterly1 |  |  |  | Irish raised to Irish and N.Irish; UK raised to UK (E\&W) and 85 Scotland |  |
|  | ALK | quarterly1quarterly1quarterly1 |  |  |  | Belgian raised to Belgian, Dutch and French |  |
| Age comp.quarterly1quarterly1quarterly1 |  |  |  |  |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. separate sex |  |
| 1979 |  |  |  |  |  |  |  |
| 1980 | Len. comp. | quarterly1quarterly1quarterly1 |  |  |  | Irish raised to Irish and N.Irish; UK raised to UK (E\&W), Sco 86 and IOM. |  |
|  | ALK | quarterly1quarterly1quarterly1 |  |  |  | Belgian raised to Belgian, Dutch and French |  |
| Age comp.quarterly1quarterly1quarterly1 |  |  |  |  |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. separate sex |  |
| 1981 |  |  |  |  |  |  |  |
| 1982 |  | As for 1980 | As for 1980 | As for 1980 |  | As for 1980, separate sex | 92 |
| 1983 |  | As for 1980 | As for 1980 | As for 1980 |  | As for 1980; sexes combined | 90 |
| 1984 | Len. comp. | quart | 2nd qtr | quarter |  | Irish raised to Irish and N.Irish 90 |  |
|  | ALK | quarter | 2nd qtr | quarterly |  | UK raised to UK (E\&W), Scotland, I.O.M., French, Dutch and Belgian |  |
|  | Age co | quarter | 2nd qtr | quarterly |  | UK + IR combined to total int. sexes combined |  |
| 1985 | Len. comp. | quarter | quarterly | quarterly |  | Irish raised to Irish and N.Irish; UK raised to UK (E\&W), Sco 92 and IOM |  |
|  | ALK | quarter | quarterly | quarterly |  | Belgian raised to Belgian, Dutch and French |  |
|  | Age co | quarter | quarterly | quarterly |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. sexes combined |  |
| 1986 | Len. comp. | quarter | quarterly | quarterly |  | Irish raised to Irish.,N.Irish and French |  |
|  | ALK | quarte | quarterly | quarterly |  | UK raised to UK (E\&W), Scotland and I.O.M.; Belgian used alone |  |
|  | Age co | . quarter | quarterly | quarterly |  | UK + Bel + IR combined to total int. |  |
| 1987 |  | As for 1986 | As for 1986 | As for 1986 |  | As for 1986 | 84 |
| 1988 |  | As for $1986$ | As for 1986 | As for 1986 |  | As for 1986 except Irish beam trawl raised using UK age comps |  |


| Year |  | Source |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| of WG | Data | UK | Belgium | Ireland | Netherlands | Derivation of international landings | $\%$ <br> sampled |
| 1989 |  | As for 1986 | As for 1986 | As for 1986 |  | As for 1986 (Irish beam trawl now sampled) | 86 |
| 1990 |  |  |  |  |  |  |  |
| 1991 |  | As for <br> 1986 | As for 1986 | As for 1986 |  | As for 1986 | 83 |
| 1992 |  | As for <br> 1986 | As for 1986 | As for 1986 |  | As for 1986 | 83 |
| 1993 |  | As for <br> 1986 | As for 1986 | As for 1986 |  | As for 1986 | 91 |
| 1994 |  | As for <br> 1986 | As for 1986 | As for 1986 |  | As for 1986 (Belgian samples supplemented with UK data) | 90 |
| 1995 |  |  |  |  |  |  |  |
| 1996 |  | As for 1986 | As for 1986 | As for 1986 | As for 1986 |  | 89 |
| 1997 |  | As for 1998 | As for 1998 | As for 1998 | As for 1998 | As for 1998 | 83 |
| 1998 | Len. comp. | quarterly quarterly quarterly quarterly |  |  |  | Irish raised to Irish., N.Irish and French; Belgian and Dutch 87 used alone |  |
|  | ALK | quarterly quarterly quarterly quarterly |  |  |  | UK raised to UK (E\&W), Scotland and I.O.M. |  |
| Age comp.quarterly quarterly quarterly quarterly |  |  |  |  |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}+\mathrm{NL}$ combined to total int. |  |
| 1999 |  | As for 1986 | As for 1986 | As for 1986 |  | As for 1986 (except UK raised to include NL landings) | 89 |
| 2000 |  | As for 1999 | As for 1999 | As for 1999 |  | As for 1999 | 88 |
| 2001 |  | As for 1998 | As for 1998 | As for 1998 | As for 1998 | As for 1998 | 87 |
| 2002 |  | As for <br> 1986 | As for 1986 | As for 1986 |  | As for 1986 | 88 |
| 2003 | Len. comp. | quarterly 1st qtr |  | quarterly |  | Belgium raised using 1st qtr values | 70 |
|  | ALK | quarter | 1 st qtr | quarterly |  | UK raised to Sco and France; Irish raised to Irish and N.Iris |  |
| Age comp.quarterly 1st qtr |  |  |  | quarterly |  | UK + Bel + IR combined to total int. |  |
| 2004 | Len. comp. | quarterly quarterly quarterly |  |  |  |  | 52 |
|  | ALK | quarter |  | quarterly |  | UK raised to Sco and France; Irish raised to Irish, N.Irish and Bel |  |
| Age comp.quarterly - |  |  |  | quarterly |  | UK + IR combined to total int. |  |
| 2005 | Len. comp. | quarterly quarterly quarterly |  |  |  |  | 81 |
|  | ALK | quarter | qrts 1,2 | quarterly |  | UK raised to Sco and France; Irish raised to Irish, N.Irish and Bel |  |


| Year |  | Source |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| of WG | Data | UK Belgium | Ireland | Netherlands | Derivation of international landings | sampled |
| Age comp.quarterly qrts 1,2 |  |  | quarterly | UK + IR combined to total int. |  |  |
| 2006 | Len. comp. | quarterly quarterly quarterly |  | 923 |  |  |
|  | ALK | quarterly quarterly | quarterly |  | UK raised to Sco and France; Irish raised to Irish, N.Irish and Bel |  |
| Age comp.quarterly quarterly quarterly |  |  |  | UK + IR combined to total int. |  |  |
| $2007$ | Len. comp. | quarterly quarterly | quarterly |  |  | 903 |
|  | ALK | quarterly quarterly | quarterly |  | UK raised to Sco and France; Irish raised to Irish and N.Irish |  |
| Age comp.quarterly quarterly quarterly |  |  |  | UK + Bel + IR combined to total int. |  |  |
| 2008 | Len. comp. | quarterly annual | quarterly |  |  | 94 |
|  | ALK | quarterly annual | quarterly |  | UK raised to Sco and France; Irish raised to Irish and N.Irish |  |
| Age comp.quarterly annual |  |  | quarterly |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. |  |
| 2009 | Len. comp. | quarterly quarterly quarterly |  |  |  | 89 |
|  | ALK | quarterly quarterly | quarterly |  | UK raised to Sco and France; Irish raised to Irish and N.Irish |  |
| Age comp.quarterly quarterly quarterly |  |  |  |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. |  |
| $2010$ | Len. comp. | quarterly quarterly quarterly |  |  | 94 |  |
|  | ALK | quarterly quarterly | quarterly |  | UK raised to Sco and France; Irish raised to Irish and N.Irish |  |
| Age comp.quarterly quarterly quarterly |  |  |  | $\mathrm{UK}+\mathrm{Bel}+\mathrm{IR}$ combined to total int. |  |  |

1 Assumed - (not explicitly stated in report)
2 Revised 2007
3 Revised 2008

Table B.2. Assessment model settings since 1991.

| Assessment Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Assessment Age Range | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ | 1-9+ |
| Fbar Age Range | 3-8 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 | 3-6 |
| Assessment Method | L.S. | L.S. | XSA | XSA | XSA | XSA | XSA | XSA | XSA | XSA | XSA | XSA | XSA | XSA |
| Tuning Fleets |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK trawl, years: ages: | $\begin{gathered} 81-90 \\ 1-8 \end{gathered}$ | $\begin{gathered} 82-91 \\ 1-8 \end{gathered}$ | $\begin{gathered} 76-92 \\ 1-8 \end{gathered}$ | $\begin{gathered} 76-93 \\ 1-8 \end{gathered}$ | $\begin{gathered} 76-94 \\ 1-8 \end{gathered}$ | - | - | - | - | - | - | - | - | - |
| UK otter, years: ages: | - | - | - | - | - | $\begin{gathered} 86-95 \\ 2-8 \end{gathered}$ | $\begin{gathered} 87-96 \\ 2-8 \end{gathered}$ | $\begin{gathered} 88-97 \\ 2-8 \end{gathered}$ | $\begin{gathered} 89-98 \\ 2-8 \end{gathered}$ | $\begin{gathered} 90-99 \\ 2-8 \end{gathered}$ | $\begin{gathered} 91-00 \\ 2-8 \end{gathered}$ | $\begin{gathered} 87-01 \\ 2-8 \end{gathered}$ | $\begin{gathered} 87-02 \\ 2-8 \end{gathered}$ | $\begin{gathered} 87-03 \\ 2-8 \end{gathered}$ |
| UK beam, years: ages: |  |  | - | - | - | - | - | - | $\begin{gathered} 89-98 \\ 2-8 \end{gathered}$ | $\begin{gathered} 90-99 \\ 2-8 \end{gathered}$ | $\begin{gathered} 91-00 \\ 2-8 \end{gathered}$ | $\begin{gathered} 89-01 \\ 2-8 \end{gathered}$ | $\begin{gathered} 89-02 \\ 2-8 \end{gathered}$ | $\begin{gathered} 89-03 \\ 2-8 \end{gathered}$ |
| Bel Beam, years: ages: | - | - | - | - | $\begin{gathered} 85-94 \\ 2-8 \end{gathered}$ | $\begin{gathered} 86-95 \\ 3-8 \end{gathered}$ | $\begin{gathered} 87-96 \\ 3-8 \end{gathered}$ | $\begin{gathered} 88-97 \\ 3-8 \end{gathered}$ | - | - | - | - | - | - |
| IR otter, years: ages: |  |  |  |  | - | - | - | - | - | - | - | $\begin{gathered} 95-01 \\ 2-8 \end{gathered}$ | $\begin{gathered} 95-02 \\ 2-8 \end{gathered}$ | $\begin{gathered} 95-03 \\ 2-8 \end{gathered}$ |
| UKBTS Sept, years: ages: |  |  | $\begin{gathered} 88-92 \\ 1-4 \end{gathered}$ | $\begin{gathered} 88-93 \\ 1-4 \end{gathered}$ | $\begin{gathered} 88-94 \\ 1-4 \end{gathered}$ | $\begin{gathered} 88-95 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-96 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-97 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-98 \\ 1-4 \end{gathered}$ | $\begin{gathered} 90-99 \\ 1-4 \end{gathered}$ | $\begin{gathered} 91-00 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-01 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-02 \\ 1-4 \end{gathered}$ | $\begin{gathered} 89-03 \\ 1-7 \end{gathered}$ |
| UKBTS Mar, years: ages: |  |  |  |  |  |  |  | $\begin{gathered} 93-97 \\ 1-4 \end{gathered}$ | $\begin{gathered} 93-98 \\ 1-4 \end{gathered}$ | $\begin{gathered} 93-99 \\ 1-4 \end{gathered}$ | $\begin{gathered} 93-99 \\ 1-4 \end{gathered}$ | $\begin{gathered} 93-99 \\ 1-4 \end{gathered}$ | $\begin{gathered} 93-99 \\ 1-4 \end{gathered}$ | - |
| IR-JPS, years: ages: |  |  |  |  |  | - | - | - | - | - | ${ }^{-}$ | $\begin{gathered} 91-01 \\ 1-6 \end{gathered}$ | $\begin{gathered} 91-02 \\ 1-6 \end{gathered}$ | - |


| Assessment Year | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Time taper |  |  | 20 yr tri | 20 yr tri | 20 yr tri | No | No | No | No | No | No | No | No | No |
| Power model ages |  |  | 1 | 0 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| P shrinkage |  |  | True | False | True | True | True | True | True | False | False | False | False | False |
| Q plateau age |  |  | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| F shrinkage S.E |  |  | 0.3 | 0.3 | 0.5 | 0.5 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 | 1.5 |
| Number of years |  |  | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| Number of ages |  |  | 5 | 5 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 |
| Fleet S.E. |  |  | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |

## Table B.2. continued.

| Assessment year |  | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Assessment model |  | ICA | ICA | ICA | ICA | ICA | ICA |
| Tuning fleets | UK(E\&W) OTB | - | - | - | - | - | - |
|  | UK(E\&W)BTS Sept | 1989-2004 | 1989-2005 | 1989-2006 | 1989-2007 | 1989-2008 | 1989-2009 |
|  | ages: | 1-7 | 2-7 | 2-7 | 2-7 | 2-7 | 2-7 |
|  | UK(E\&W)BTS March | - | - | - | - | - | - |
|  | UK(E\&W)BT | - | - | - | - | - | - |
|  | IR-OTB | - | - | - | - | - | - |
|  | UK(NI) GFS Mar | 1992-2004 | 1992-2005 | 1992-2006 | 1992-2007 | 1992-2008 | 1992-2009 |
|  | Biomass index |  |  |  |  |  |  |
|  | UK(NI) GFS Oct | 1992-2004 | 1992-2005 | 1992-2006 | 1992-2007 | 1992-2008 | 1992-2009 |
|  | Biomass index |  |  |  |  |  |  |
| Time series weights |  | Full time-series - unweighted | Full time-series - unweighted | Full time-series - unweighted | Full time-series - unweighted | Full time-series - unweighted | Full time-series - unweighted |
| Num years for separable |  | 5 | 5 | 6 | 7 | 8 | 9 |
| Reference age |  | 4 | 5 | 5 | 5 | 5 | 5 |
| Terminal S |  | 1 | 1 | 1 | 1 | 1 | 1 |
| Catchability model fitted |  | Linear | Linear | Linear | Linear | Linear | Linear |
| SRR fitted |  | No | No | No | No | No | No |
| Landings number-at-age, range: |  | 1-9+ | 2-9+ | 2-9+ | 2-9+ | 2-9+ | 2-9+ |

## Stock Annex 6.8: Sole in VIIa

- Stock Annex 6.8 Sole VIIa: for latest update see WGCSE 2009, Annex 06.8 Sole VIIa


## Stock Annex 7.2: Cod in VIIe-k

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Cod in VIIe-k (Celtic Sea cod) |
| :--- | :--- |
| Expert Group | Celtic Sea Working Group |
| Date | WKROUND 2009, WGCSE 2009-2010 |
| Revised by | Robert Bellail, Lionel Pawlowski |

## A. General

## A.1. Stock definition

Since 1997, this assessment has related to the cod in Divisions VIIe-k, covering the Western Channel and the Celtic Sea. The assessed area has gradually increased from VIIfg before 1994 to VIIfgh, to VIIefgh in 1996 and finally to VIIe-k.

Up to 2008, the management area was set in Divisions VIIb-k,VIII, IX, X, and CECAF 34.1.1 which does not correspond to the area assessed.

In 1994, at the request of ACFM, the ICES Working Group on Southern Shelf Demersal Stocks (WGSSDS) studied the possible extension of the area assessed from VIIfg to VIIfgh. Examination of data from surveys and logbooks indicated a continuity of the distribution of VIIg cod into VIIh. Depending on the year, catches in Division VIIh represented $9-15 \%$ of the catches in VIIfg, with a coincidence of years of peak or low catches in both areas. Therefore, catches from VIIh were included in the assessment. In 1996, at the request of ACFM, WGSSDS studied the possible extension of the area assessed from VIIfgh to VIIefgh. The population dynamics parameters for VIIfgh and VIIe cod were examined and compared for the period 1988-1994, when independent tuning fleets, international catch-at-age, mean weights-at-age in the landings and in the stocks were available for both areas. Patterns of F were consistent between VIIe and VII fgh in earlier years (1988-1990), and SSBs trends were similar in the period 1988-1992. The patterns of recruitment (age 1) were found to be fairly consistent through this period 1988-1994, though it cannot be assumed that this consistency was also valid in earlier years when catch-at-age were only available in Divisions VIIf, g, h. It was therefore decided to combine Western Channel Cod with the Celtic Sea Cod assessment for the years 1988-1995, but an independent assessment of Celtic sea Cod in VIIfgh was maintained for the longer period available 1971-1995. This was to allow scaling of the historical (1971-1987) SSBs and recruitments values from VIIfgh to VIIe-h.

At WGSSDS 1997, due to the lack of a long independent series of catch-at-age in Divisions VIIj,k, the estimate of landings from Divisions VIIjk was discussed and it was decided to combine the data of Divisions VIIe,f,g,h and Divisions VIIjk for the period 1993-1996 and to raise the data in Divisions VIIe-h to landings in Divisions VIIe-k for the period 1988-1992. The results of an XSA assessment of this series in Divisions VIIe-k for 1988-1996 had been compared with the results of the assessment in Divisions VIIe-h in terms of trends of F, SSB and recruitment. Patterns of these parameters were found very similar and the merging of Divisions VIIjk with Divisions VIIe$h$ mainly resulted in a scaling upwards of SSB and recruitment. The new assessment areas comprised cod in Divisions VIIe-k.

At the 1999 WGSSDS meeting, an alternative procedure to the tedious re-scaling of SSB and recruitment of the earlier series 1971-1987 in VIIfgh to VIIe-k every year was proposed (Bellail, 1999, WD3). A long series of landings data from 1971-1987 was reconstructed. An average raising factor (1.24) from VIIfgh to VIIe-k in the period 1988-1997 was applied to VIIfgh landings of the series 1971-1987. Results of assessment in terms of SSB and R were very close to those obtained when these parameters were scaled. ACFM accepted this procedure.

In the past, few biological criteria have been used to justify the widening the stock area. However, recent tagging work by Ireland and the UK supports the idea that there is a resident stock in the Celtic Sea and Western Channel (VIIe-k) and mixing with other areas appears to be minimal. The Irish Sea front, running from SE Ireland (Carnsore point) to the Welsh Coast, appears to act as boundary between the Irish Sea and Celtic Sea stock. Juveniles found close to the SE Irish Coast (south of VIIa) are considered part of the Celtic Sea stock.

Migrations are known to occur in this cod stock. Cod can be caught throughout the English Channel (ICES Areas VIId and VIIe) in autumn (quarter 4) and winter (quarter 1), being more aggregated during the spawning season in January/February. Electronic tagging experiments in the English Channel (VIId and VIIe) have shown that cod tagged on or close to English Channel spawning grounds in quarters 4 and 1 either remain close to the point of release (residency), or move to feeding grounds to the south and/or west. Smaller fish $(<50 \mathrm{~cm})$ are more likely to be resident. Migrants tend to move offshore to deeper areas, whereas the habitat selection of residents is less clearcut.

From the migratory phenotypes identified by electronic tagging, historical markrecapture experiments can be re-evaluated. Although sample size is limited, results from data on the movements of adult cod ( $>50 \mathrm{~cm}$ ) show that, after tagging in VIIe (the western Channel) in quarters 1 and $4,47 \%$ of $\operatorname{cod}(27$ of 58$)$ are recaptured in ICES Areas VIIf through VIIj, while $48 \%$ are recaptured in VIIe (i.e. are probably resident). In contrast, no adult cod tagged in VIId were recaptured in ICES Areas VIIf through VIIj, $5 \%$ moved into VIIe and $51 \%$ remain in VIId. Juvenile cod are more likly to be recaptured in the same area that they were tagged in. These figures vary slightly when recaptures are separated into autumn/winter and spring/summer seasons, but are broadly comparable. The data therefore provide evidence that cod in the eastern English Channel and western English Channel might be classed as separate substocks, and that movement of cod between eastern English Channel and the Celtic Sea is limited, whereas movement between the western Channel and the Celtic Sea is frequent.

## A.2. Fishery

Cod in Divisions VIIe-k are mainly taken as components of catches in mixed demersal trawl fisheries with a minor part by gillnets. Landings are made throughout the year but are generally more abundant during the first semester. Constraining TACs set since 2003 and the impact of the Trevose Head Closure applied since 2005 have led the landings to spread across the first three quarters of the year.

WGSSDS has been collating a database of landings and effort for the Celtic Sea. Available data on cod landings are analyzed and presented. Effort data is not yet fully available for similar investigations. Recent temporal and spatial patterns in landings distributions for the main fleets catching Celtic Sea Cod are shown in Figure A.2.1 and Figure A.2.2. Highest landings are in quarter 1 when the cod aggregate to spawn. There is an indication that Q1 landings have declined in 2006 and 2007 as a
result of the closure of a known spawning area at Trevose Head, although this was not the case in 2005 the first year of introduction of the closure. In most years there is a distinct peak in landings in February or March. The scale of this peak may be related to the relative strength of age 2 fish entering the fishery. The majority of the landings come from VIIg, $\sim 55 \%$, and the relative contributions of different ICES Divisions to the landings has been fairly stable over recent years. In 2002 there were larger than normal landings from rectangle 30E4 in VIIf.

The majority of the landings are made by demersal trawls targeting roundfish (i.e. cod, haddock and whiting), although, in recent years an increasing component have been from gillnets and otter trawls targeting Nephrops and benthic species.

## A.3. Ecosystem aspects

No environmental drivers are known for this stock.

## B. Data

## B.1. Commercial catch

## Landings

On a quarterly basis, France and UK (E+W) have provided catch numbers-at-age and catch weights-at-age for their landings. Ireland has provided with the same data in Divisions VIIg and j separately and estimates of misreporting in VIIg. Landings only are available for Belgium.

Irish data are first aggregated to the landings in VIIe- k and then both datasets for France, UK and Ireland are added and raised to international landings taking into account Belgian data. Then the quarterly datasets are summed up to the annual values.

As a consequence of an update to the French database of landings statistics, some minor revisions (downward) have been applied since 2002 and the updated datasets for international landings.

Nothing is hidden in the aggregating procedure but the level of available data has changed and consequently the aggregation procedures. Compiling the previous reports of the WGSSDS and before the reports of the WGIRCS shows the following datasets available and the history of the aggregation procedures to produce the landings numbers-at-age series:

| Year range | Landings VIIe-k | Length structure (Ls) VIIe-k | Age structure (As) VIle-k |
| :---: | :---: | :---: | :---: |
| 1971-1976 | Annual VIIfgh expanded to Annual VIIe-k using the mean landings VIIe-k 1988-1997 over the mean | UK VII fg raised to international landings in VIIfg | UK alks VIIa to UK Ls VIIfg then UK VIIfg As raised to international landings |
| 1977-1980 | landings VIIfgh 1988- <br> 1997 as a ratio | UK VIIfg + FRVIIfg raised to international landings in VIIfg | UK alks VIIa to UK Ls VIIfg and FR Ls VIIfg then As summed and raised to international landings |
| 1981-1987 |  | UK VIIfg <br> FR VIIfg raised to VIIfgh | FR alks VIIfg to UK\&FR <br> Ls VIIfg then As summed and raised to international landings |


| Year range Landings VIIe-k | Length structure (Ls) VIle-k | Age structure (As) VIle-k |
| :---: | :---: | :---: |
| 1988-1989 | UK VIIfg <br> UK VIIe <br> FR VIIfg raised to VIIfgh | FR alks VIIfg to FR Ls UK alks VIIfg to UK Ls UK alks VIIe to UK Ls then As summed and raised to international landings |
| 1990 | UK VIIfg <br> UK VIIe <br> FR VIIfg raised to VIIfgh IR VIIg | FR alks VIIfg to FR Ls UK alks VIIfg to UK Ls UK alks VIIe to UK Ls IR alks VIIg to IR Ls then As summed and raised to international landings |
| 1991-1998 | UK VIIfg <br> UK VIIe <br> FR VIIfg raised to VIIfgh <br> IR VIIg <br> IR VIIj annual | FR alks VIIfg to FR Ls UK alks VIIfg to UK Ls UK alks VIIe to UK Ls IR alk VIIg to IR VIIg Ls IR alk VIIj to IR VIIj Ls then sum of As VIIg or VIIfg raised to VIIfgh international, <br> As UK VIIe raised to VIIe international, <br> As IR VIIj raised to VIIjk international landings, (VIIfgh internat+ VIIe internat + VIIjk internat) = VIIek |
| 1999-2001 | UK VIIfg <br> UK VIIe <br> FR VIIfg raised to VIIfgh <br> IR VIIg <br> IR VIIj quarterly | FR alks VIIfgh to FR Ls UK alks VIIfg to UK Ls UK alks VIIe to UK Ls IR alk VIIg to IR VIIg Ls IR alk VIIj to IR VIIj Ls then sum of As VIIg or VIIfg raised to VIIfgh international, <br> As UK VIIe raised to VIIe international, <br> As IR VIIj raised to VIIjk international landings, (VIIfgh internat+ VIIe internat + VIIjk internat) $=$ VIIek |


| Year range | Landings VIIe-k | Length structure (Ls) VIIe-k |
| :--- | :--- | :--- |
| $2002-\ldots$ | FR-VIIe-k | Age structure (As) VIIe-k |
|  | UK VIIe-k | FR alks VIIfgh to FR Ls |
|  | IR VIIg | UK alks VIIfg to UK Ls |
|  | IR VIIj | UK alks VIIe to UK Ls |
|  |  | IR alk VIIg to IR VIIg Ls |
|  |  | IR alk VIIj to IR VIIj Ls |
|  |  | Then sum As UK raised |
|  |  | to UK landings in VIIe-k, |
|  |  | Sum As IR raised to IR |
|  |  | landings in VIIe-k, |
|  |  | Then AsUK+As IR+ As FR |
|  |  | raised to international |
|  |  | landings |

At each step of the aggregations, mean weight-at-age is the weighted mean by num-bers-at-age.

## Discards

Discards data sampled under EU/DCR since 2003 have been generally presented in previous WGSSDS but not used in the assessments as they do not cover all the main fleets and quarters yet.

Due to the annual management system adopted by the French POs since 2003 in response to the quota restrictions, high grading has occurred in the French fishery, mainly in VIIfgh. On an annual basis, a procedure using both the UK and French landings length data enabled estimation of the French high grading for the years 2003-2005 (WD 1, WGSSDS 2006). The adjustments were reapplied to improve estimates of French landings from 2003 at the ICES WKROUND 2009. This procedure could not be used in later years as high-grading has also occurred in that years.

In 2008 the French self-sampling programme on Celtic Sea cod has produced datasets enabling estimation of discarding and high-grading rates on a quarterly basis. Assuming the same pattern of discarding in recent years, estimates of French discarding and high-grading back to 2006 were also computed. Estimates of high-grading were also calculated for the French tuning fleets used in the analysis (ICES WKROUND, 2009, WD 17). Since the WKROUND, the database of the 2008 self sampling has increased and led to a slight update of the estimates of the level of French high-grading.

## Lpue

The table below summarizes the available data.

| Name | Area |  |
| :--- | :---: | :--- |
| FR gadoid fleet $^{1}$ | VIIfgh | series |
| FR Nephrops fleet $^{1}$ | VIIfgh | $1983-\ldots$ |
| FR otter trawlers $^{2}$ | VIIe | $1983-\ldots$ |
| FR otter trawlers $^{2}$ | VIIfgh | $1983-\ldots$ |
| FR otter trawlers $^{2}$ | VIIe-k | $1983-\ldots$ |
| UK otter trawlers $^{\text {UK otter trawlers }}$ | VIIe | $1983-\ldots$ |
| UK beam trawlers | VIIe-k | $1972-\ldots$ |
| IR otter trawlers $_{\text {IR beam trawlers }}^{\text {IR Scottish seiners }}$ | VIIe-k | $1972-\ldots$ |
| IR otter trawlers | VIIg | $1978-\ldots$ |
| IR beam trawlers | VIIg | $1995-\ldots$ |
| IR Scottish seiners | VIIg | $1995-\ldots$ |

${ }^{1}$ For Q2+3+4 for consistency with the Trevose Head Closure since 2005 during the first quarter.
${ }^{2}$ Annual values, including the Fr gadoid and Nephrops fleets.

## B.2. Biological

## Weights-at-age

At the 1999 WGSSDS, data for the years 1971-1980 were set to the average 1981-1997. A revision was carried out at 2001 WGSSDS where the values for the period 19711980 were set to the average values for 1981-2000. Depending on the annual datasets available by country for the period 1988-2001, catch weights-at-age data were calculated as the weighted means from French, Irish and UK datasets. Since 2002, VIIe-k catch weights-at-age have been calculated as the annual weighted means of French, Irish and UK datasets in VIIe-k.

## Maturity

The maturity ogive applied since 1999, was estimated from the datasets of the UKWCGFS survey (first quarter) has been used for the overall series. It replaced an assumed ogive used for the year prior to 1999, derived from Irish Sea cod data, when both stocks (VIIa and VIIfg) were assessed in the Irish Sea and Bristol Channel WG up to 1992. Table below summarizes the maturity ogives used.

| Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5 +}$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Before 1999 | 0.00 | 0.05 | 1.00 | 1.00 | 1.00 |
| Current | 0.00 | 0.39 | 0.87 | 0.93 | 1.00 |

## Natural mortality

In the assessments, natural mortality is assumed to be constant $(\mathrm{M}=0.2)$ for the whole range of years and ages.

## B.3. Surveys

Three survey-series are available. The common range of ages used is $1-5$ :
The discontinued UK-WCGFS (1986-2004), conducted during the first quarter, is generally truncated into a shorter series (1992-2004) as it showed a strong trend (dome-shaped) when using the full series. This pattern is related to the progressive extension of the studied area of this survey from VIIe to VIIefgh over the years. This time-series only contributes to the estimates at older ages ( 4 and older). Due to the lack of new data the series is no longer used for calibration.

The FR-EVHOE survey (1997-...), during the fourth quarter, covers the Divisions VIIfghj. The full series is used.

The IrGFS survey (2003-...), during the fourth quarter, in VIIg and VIIj is also used in the assessment. It is the main contributor to the terminal year estimates, partly because this series is short.

The absolute numbers of cods caught in all of these surveys are extremely low.

## B.4. Commercial cpue

Two French commercial fleets are used for tuning: the French trawlers targeting Gadoids in Divisions VIIf, g, h (FR-GADOIDS) and the French Nephrops trawlers in VIIf,g,h (FR-NEPHROPS), for which cod is generally a bycatch. Both fleets account on average for $\sim 30 \%$ of the international landings from 1988; the series starts in 1983. Other commercial fleets used are the English West Coast otter trawlers (UK-WECOT) in VIIe from 1988 and the Irish 7J otter trawlers (IR-7J-OT) in VIIj from 1995. Both fleets fish throughout the majority of the assessed area.

## B.5. Other relevant data

Input from industry
No new datasets.

## C. Historical stock development

Model used:
The Separable VPA was used at the former Irish Sea and Bristol Channel WG and the Laurec-Shepherd model in the period 1987-1992. The XSA was the model used subsequently. SURBA was also used for survey catch-at-age analysis in 2005-2007.

Corrections for some misreporting estimates have been integrated into the datasets used in the assessment but the change of discarding practices to manage the restricting national quotas may impact the assessment. This also affects the reliability of the commercial tuning fleets used.

In previous assessments (2006, 2007 and 2008), adding a new year of data has generally raised the stock numbers at younger ages (age 1 and 2) resulting in increased estimates of recruitment strength. These upwards revisions are considered a result of the recent high-grading practices. Given this uncertainty and the recent reports from the industry of underreporting the XSA assessment, which assumes unbiased catch data cannot be applied. Improved datasets on landings, recorded and high-grading are required before XSA could be used.

WKROUND (2009) evaluated XSA with adjusted recent catch levels against B-Adapt and the SAM state-space model, which estimate additional unallocated mortality. All models exhibited different patterns in the recent years with a high degree of uncertainty. The Group concluded that no model could be recommended as a basis for providing advice on recent stock trends until further investigations or additional datasets were available to resolve the situation.

## D. Short-term projection

No decision has been taken on the forecast methodology.

## E. Sensitivity analysis and medium-term projections

Medium-term forecasts are not provided for this stock.

## F. Long-term projections

Long-term forecasts are not provided for this stock.

## G. Biological reference points

Reference points


Due to the current uncertainties on the state of this stock, the Benchmark WK is unable to make new proposals for the Reference Points and the 2004 values remain.

## H. Other issues

None.

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Figure A.2.1. Temporal and spatial patterns in landings patterns for Celtic Sea cod (VIIe-k).


Figure A.2.2. The spatial and temporal distribution of cod landings from the Celtic Sea, from 2000-2007 by gear type. The closed rectangles are highlighted in yellow. Each year is scaled to, the maximum.


Figure A.2.2. continued.


Figure A.2.2. continued.


Figure A.2.2. continued.


Figure A.2.2. continued.

## Stock Annex 7.4: Haddock in VIIb-k

- Stock Annex 7.4 HaddockVIIb-k for latest update see WGCSE 2009, Annex 07.4 Haddock VIIb-k


## Stock Annex 7.5: FU17, Aran Grounds

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Aran Grounds Nephrops (FU17) |
| :--- | :--- |
| Date | 06 March 2009 (WKNEPH 2009) |
| Revised by | Colm Lordan and Jennifer Doyle (WKNEPH, 2009) |

## A. General

## A.1. Stock definition

Nephrops is limited to muddy habitat, and requires sediment with a silt and clay content of between $10-100 \%$ to excavate its burrows, and this means that the distribution of suitable sediment defines the species distribution. Adult Nephrops probably only undertake very small scale movements (a few 100 m ) but larval transfer may occur between separate mud patches in some areas. In FU17, the main Nephrops stock inhabits an extensive area of muddy sediment known as the Aran Grounds which lie to the west and southwest of the Aran Islands, there are also smaller discrete mud patches in Galway Bay and Slyne Head.

## A.2. Fishery

In recent years the Nephrops stock in FU17 are almost exclusively exploited by Irish vessels. Figure A.2.1 shows the spatial distribution of landings and lpue for Irish otter trawl vessels in 2005 using logbook and VMS data linked together to give finer spatial resolution. The Aran ground fishery is clearly highlighted.

The Nephrops fishery 'at the back of the Aran Islands' can be considered the mainstay of the Ros a Mhíl fleet. Without this Nephrops fishery the majority of vessels in the fleet would cease being economically viable (Meredith, 1999). The Irish fishery consists of entirely of otter trawl vessels. The majority of vessels use twin-rigs and 80 mm . Smaller vessels do use 70 mm with a SMP. Some vessels have using 90 mm . Vessels from Ros a Mhíl, Dingle, Union Hall, Dunmore East, Clogherhead and Kinsale mainly exploit the fishery.

The number of Irish vessels reporting Nephrops landings from FU17 has fluctuated around $50 / \mathrm{yr}$ (Figure A.2.2). Around 18 vessels report landings in excess of 10 t . These are the main vessels in the fishery accounting for around $85 \%$ of the total landings. The majority of these vessels are between 20-22 m overall length (Figure A.2.3). There has been a slight shift to lager vessels over time. The majority of vessels are in the power range of 200-400 KW (Figure A.2.4). There has also been a shift to more powerful vessels over time with the introduction of twin-rigs to the fishery in the early 2000s. Most of the larger boats move freely between the Nephrops fisheries in FUs $15,16,20-22$ and other areas depending on the tides and weather.

The fishery shows a distinctive seasonal pattern with highest landings, catches, lpue and cpue in April-June and October-November. The monthly landings time-series with the average pattern is shown in Figure A.2.5. The first period of elevated landings is associated with the emergence of females from their burrows post hatching of their eggs. The sex ratio during this period is biased towards females (Figure A.2.6). Females mature quickly during the early summer and spawning occurs in July and August. This is coincident with a decline in landings and cpue in the fishery. The

Ros a Mhíl fleet traditionally tie up in August each year for maintenance and refurbishment.

The following TCMs are in place for Nephrops in VII (excluding VIIa) after EC 850/98: Minimum Landing Sizes (MLS); total length $>85 \mathrm{~mm}$, carapace length $>25 \mathrm{~mm}$, tail length $>46 \mathrm{~mm}$. Mesh Size Restrictions; Vessels targeting Nephrops using towed gears having at least $35 \%$ by weight of this species on board will require 70 mm diamond mesh plus an 80 mm square mesh panel as a minimum or having at least $30 \%$ by weight of Nephrops on board will require $80-99 \mathrm{~mm}$ diamond mesh.

## A.3. Ecosystem aspects

## Physical oceanography

The Aran Ground is coincident with a pool of oceanic water, which is rich in nutrients and low in dissolved oxygen. The currents throughout the water column over the ground are generally weak although there is a well-documented bottom density front on the eastern flank of the ground (Nolan and Lyons, 2006). This is a seasonal feature, which establishes in May and persists until autumn. The front causes a persistent jet like flow from south to north close to the seabed through the Nephrops ground. The mean position of jet varies from year to year by up to 30 km . Timing and position of the jet may influence recruitment and settlement success of postlarval Nephrops since it could potentially advect larval from the area. Salinity differences, due to over winter fresh water input, are thought to heavily influence the density structure and location of this front. Until a time-series of recruitment and jet dynamics is established it is not possible to draw any firm conclusions about the impact of this ecosystem feature on the stock and fishery. Potential sinks for advected larvae include Slyne head and possibly Galway Bay.

## Temperature and salinity time-series

An emerging time-series of temperature and salinity data are available for a transect through the Aran Grounds (Nolan and Lyons, 2006). In all years since 1999 (except 2001) the $53^{\circ} \mathrm{N}$ section has exhibited positive anomalies in temperature of between $0.2^{\circ} \mathrm{C}$ and $2^{\circ} \mathrm{C}$ (Figure A.3.1). In 2001, the temperature anomaly from the long-term climatology was zero. Years with lower temperature anomalies seem to coincide with years of strongly negative salinity anomalies (e.g. 2001 and 2005, 2006) perhaps reflecting the limited influence of ENAW on the section in those years as the section is dominated by coastal discharges from the Loire and Shannon. Salinity anomalies along $53^{\circ} \mathrm{N}$ range from -0.3 to +0.1 psu over the period. The freshest years were 2001, 2005 and 2006. In 2000, 2003 and 2004 ENAW has a stronger influence on the salinity structure and positive anomalies in salinity from the long-term climatology are the result. The higher UWTV abundance in 2003 and 2004 is coincident with the warmest anomaly but the time-series remains too short to draw definitive conclusions.

## Sediment distribution

There is a growing body of information on the spatial extent of the sediment suitable for Nephrops from UWTV surveys, seabed mapping programmes and the fishing industry. Figure A.3.1 depicts contour and post plots of the a) mean size (phi) and classification based on the Friedman and Sanders (1978) scales and b) sorting ( $\sigma \mathrm{g}$ ) of the sediments on the Aran Grounds based on PSA results from samples collected from 2002-2006 UWTV surveys. The majority of the ground has similar mean particle size at around $4-5 \mu \mathrm{~m}$. There are some patches of softer silt towards the middle of the ground. Figure A.3.2 is bathymetry of the Aran grounds obtained from seabed map-
ping programmes. The eastern flank of the ground shallows up quickly but the majority of the ground is gradually deepening from around 100 m to 110 m with the deepest parts to the southwest.

## B. Data

The table below summarises the available data for this stock and attempts to quantify the quality subjectively.


## B.1. Commercial catch

Prior to 1988 landings data for this fishery are only available to the WG for France. Since 1988 reported landings data for the Irish fleet were obtained from EU logbooks. The quality of landings data is not well known. In earlier, years there are no landings from Ireland although there was probably some catch. The Irish landings have been close to quota for this TAC area since around 1997 (Figure B.1.1). In more recent years (2003-2005 and 2008) there are a few observations of both under and over reporting but it is not possible to correct landings using these as it is not known how representative they might be.

Landings length and sex compositions were estimated from port sampling by Ireland (between 1995-2001). There was a perception during this period that that discarding was not significant. In 2002 a new catch self-sampling programme was put in place. This involves unsorted catch and discard samples being provided by vessels or collected by observers at sea on discard trips. The catch sample is partitioned into landings and discards using an onboard discard selection ogive derived for the discard samples (Table B.1.1). Sampling effort is stratified monthly but quarterly aggregations are used to derive length distributions and selection ogives. The length-weight regression parameters given in Table B.2.1 are used to calculate sampled weights and appropriate quarterly raising factors. The sampling intensity and coverage has varied over the time-series (Table B.1.1). The quality of the sampling has not yet been qualitatively assessed in terms of precision and accuracy.

Nephrops landings and discards from the Aran Grounds have not been sampled for the majority of 2006 and all 2007 due to a lack of co-operation by the industry. However, sampling resumed in 2008 and the intensity and coverage is considered the best to date.

Fish and other bycatches in the fishery have been collected by on board observers since 1994. The number of trips is variable over time with a gap in the series in 2006 and 2007.

## B.2. Biological

Biological parameters for this stock are outlined in Table B.2.1.

## Length-weight

Mean weights-at-age for this stock are estimated from studies on Scottish stocks by Pope and Thomas (1955). This relationship was examined in 2003 and it seemed appropriate. Given the variability in length-weight parameters found in Allan et al., 2009 it would be worth monitoring these more closely in the future.

## Natural mortality

A natural mortality rate of 0.3 was assumed for all age classes and years for males and immature females, with a value of 0.2 for mature females. The lower value for mature females reflects the reduced burrow emergence while ovigerous and hence an assumed reduction in predation. The accuracy of these assumptions is unknown. Cod are not common on the Aran Grounds but other potential predators include dogfish, monkfish megrim and gurnards. Stomach contents data on the Irish GFS could be used to examine this in the future.

## Maturity

The $L_{50}$ of females using a macroscopic visual maturity scale is known to vary depending sampling month (Lordan and Gerritsen, 2006). The L50 in July was chosen as the most appropriate estimate given the maturity schedules observed (Figure B.2.1). It is worth mentioning that commercial vessel surveys in November 2001 and in June 2002 demonstrated considerable differences between the maturity schedules of female Nephrops sampled in shallower waters of Galway Bay compared with the Aran Grounds.

Proportion of F and M prior to spawning was specified as zero to give estimates of spawning stock biomass at January 1. In the absence of independent estimates, the mean weights-at-age in the total catch were assumed to represent the mean weights in the stock.

## Discard survival

Given the trip durations ( $\sim 5$ days average) and behaviour of the fleet the majority of discards on the Aran Grounds are returned to the sea over suitable sediment. The proportion scavenged by birds is probably quite low. Tow durations, volume of catches, prolonged sorting on deck and relatively high density of Nephrops on the seabed probably results in relatively low discard survival. This is estimated to be around $10 \%$.

## B.3. Surveys

Since 2002 Ireland has conducted underwater television survey (UWTV) annually on the main Nephrops grounds - Aran grounds. Indicator camera stations are also carried out on the adjacent grounds of Galway Bay and Slyne Head weather and time permitting. The surveys were based on a randomised fixed grid design. The methods used during the survey were similar to those employed for UWTV surveys of Nephrops stocks around Scotland and elsewhere and are documented by WKNEPHTV (ICES, 2007).

A number of factors are suspected to contribute bias to the surveys. In order to use the survey abundance estimate as an absolute it is necessary to correct for these po-
tential biases. The history of bias estimates are given in the following table and are based on simulation models, preliminary experimentation and expert opinion, the biases associated with the estimates of Nephrops abundance in the Aran Grounds are:

|  |  | detection <br>  <br>  <br> Time period |  | species <br> identification occupancy bias |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| FU17: Aran | $<=2009$ | 1.35 | 0.9 | 1.05 | 1 | 1.3 |

## B.4. Commercial cpue

Prior to 1988 landings data for this fishery are only available to the WG for France. Since 1988 reported landings data for the Irish fleet were obtained from EU logbooks (Table B.4.1).

Effort data for FU17 is available from 1995 for the Irish otter trawl Nephrops directed fleet (Table B.4.2). A threshold of $30 \%$ of Nephrops in reported landings by trip is used to identify the catches and effort of this fleet. This threshold was based on an analysis of the trip-by-trip catch compositions. In 2007 this fleet accounted for $\sim 90 \%$ of the landings and compared with an average of $70 \%$ over the time period. These data have not been standardised to take into account vessel or efficiency changes during the time period. Landings per unit of effort (lpues) have been fluctuating around an average of $39 \mathrm{~kg} / \mathrm{hr}$ with an increasing trend since 2004, to the highest observed ( $59 \mathrm{~kg} / \mathrm{hr}$ ) in the time-series in 2007 (Figure B.4.1).

## B.5. Other relevant data

## C. Historical stock development

Age structured XSA assessment for this stock was carried Nephrops WG in 2003 (ICES, 2003). The results were considered unreliable for several reasons most importantly; inadequate historical sampling of catch, growth and natural mortality assumptions and concern about accuracy of tuning data. Since then the focus has been on developing a time-series of UWTV survey data as the basis of assessment and advice for this stock.

The 2009 Benchmark decided on the following procedure:
1 ) Survey indices are worked up annually resulting in the TV index.
2 ) Adjust index for bias (see Section B.3). The combined effect of these biases is to be applied to the new survey index.
3 ) Generate mean weight in landings. Check the time-series of mean landing weights for evidence of a trend in the most recent period. If there is no firm evidence of a recent trend in mean weight use the average of the three most recent years. If, however, there is strong evidence of a recent trend then apply most recent value (don't attempt to extrapolate the trend further in the future).

## D. Short-term projection

1 ) The catch option table will include the harvest ratios associated with fishing at $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max. }}$. These values have been estimated by the Benchmark Workshop (see Section 9.2) and are to be revisited by subsequent bench-
mark groups. The values are FU specific and have been put in the Stock Annexes.

2 ) Create catch option table on the basis of a range of harvest ratios ranging from 0 to the maximum observed ratio or the ratio equating to $\mathrm{F}_{\text {max, }}$ whichever is the larger. Insert the harvest ratios from step 4 and also the current harvest ratio.
3 ) Multiply the survey index by the harvest ratios to give the number of total removals.

4 ) Create a landings number by applying a discard factor. This conversion factor has been estimated by the Benchmark Workshop and is to be revisited at subsequent benchmark groups. The value is FU specific and has been put in the Stock Annex.
5 ) Produce landings biomass by applying mean weight.
The suggested catch option table format is as follows.

|  | Implied fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Harvest rate | Survey Index | Retained number | Landings (tonnes) |
|  | 0\% | 12345 | 0 | 0.00 |
|  | 2\% | " | 247 | 123.45 |
|  | 4\% | " | 494 | 246.90 |
|  | 6\% | " | 741 | 370.35 |
|  | 8\% | " | 988 | 493.80 |
| $\mathrm{F}_{0.1}$ | 8.60\% | " | 1062 | 530.84 |
|  | 10\% | " | 1235 | 617.25 |
|  | 12\% | " | 1481 | 740.70 |
| $\mathrm{F}_{\text {max }}$ | 13.50\% | " | 1667 | 833.29 |
|  | 14\% | " | 1728 | 864.15 |
|  | 16\% | " | 1975 | 987.60 |
|  | 18\% | " | 2222 | 1111.05 |
|  | 20\% | " | 2469 | 1234.50 |
|  | 22\% | " | 2716 | 1357.95 |
| Fcurrent | 21.5\% | " | 2654 | 1327.09 |

## E. Medium-term projections

None presented.

## F. Long-term projections

None presented.

## G. Biological reference points

The time-series of available length frequencies were insufficient to generate reliable estimates of $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max }}$.

## H. Other issues

## I. References

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Nolan,G.D. and Lyons. 2006. Ocean climate variability on the western Irish Shelf, an emerging time series., K., Proceedings of the ICES Annual Science Conference, Theme Session C, C:28.

Table B.1.1. Nephrops in FU17 (Aran Grounds) Landings and discard numbers by year and sex.

|  | Female Numbers '000s |  | Male Numbers '000s |  | Both sexes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Discards | Landings | Discards | \% Discard |
| 2001 | 18,665 | 12,161 | 29,949 | 13,250 | $34 \%$ |
| 2002 | 23,105 | 9,374 | 31,256 | 8,326 | $25 \%$ |
| 2003 | 14,530 | 9,577 | 29,538 | 8,744 | $29 \%$ |
| 2004 | 16,109 | 7,068 | 12,930 | 4,282 | $28 \%$ |
| 2005 | 20,280 | 11,383 | 21,828 | 8,967 | $33 \%$ |
| 2006 | No Sampling |  |  |  |  |
| 2007 |  |  |  |  |  |

Table B.2.2. Numbers of samples and numbers measured for the FU17 Nephrops Stock by year.

| Number of Samples |  |  | Total numbers of Nephrops measured |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Graded |  |  |  | Graded |  |  |
| Year | Landings | Catch | Discards | Year | Landings | Catch | Discards |
| 1990 | 24 |  |  | 1990 | 10451 |  |  |
| 1991 | 20 |  |  | 1991 | 8260 |  |  |
| 1992 | 0 |  |  | 1992 | 0 |  |  |
| 1993 | 0 |  |  | 1993 | 0 |  |  |
| 1994 | 0 |  |  | 1994 | 0 |  |  |
| 1995 | 13 |  |  | 1995 | 6370 |  |  |
| 1996 | 3 |  |  | 1996 | 1440 |  |  |
| 1997 | 11 |  |  | 1997 | 5203 |  |  |
| 1998 | 12 |  |  | 1998 | 5388 |  |  |
| 1999 | 16 |  |  | 1999 | 6944 |  |  |
| 2000 | 5 |  |  | 2000 | 2255 |  |  |
| 2001 | 32 | 5 | 5 | 2001 | 13231 | 3194 | 3891 |
| 2002 |  | 13 |  | 2002 |  | 9399 |  |
| 2003 | 1 | 9 | 9 | 2003 |  | 6284 | 4829 |
| 2004 |  | 14 | 14 | 2004 | 578 | 12934 | 13167 |
| 2005 |  | 13 | 9 | 2005 |  | 8729 | 7559 |
| 2006 |  | 2 | 0 | 2006 |  | 767 | 436 |
| 2007 |  | 0 | 0 | 2007 |  |  |  |
| 2008 |  | 19 | 18 | 2008 |  | 4944 | 8701 |

Table B.2.1. Biological Input Parameters for FU17 Nephrops Stock.

| Parameter | Value | Source |
| :--- | :---: | :--- |
| Discard Survival | $10 \%$ | WKNEPH 2009 |
| MALES |  |  |
| Growth - K | 0.150 | based on FU15 and FU16 |
| Growth - L(inf) | 60 | based on FU15 |
| Natural mortality - M | 0.3 | assumed, in line with other stocks |
| Length/weight - a | 0.000322 | based on Scottish data (Pope and Thomas, 1955) |
| Length/weight - b | 3.207 | $"$ |
| FEMALES |  |  |
| Immature Growth | 0.150 | based on FU15 and FU16 |
| Growth - K | 60 | based on FU15 |
| Growth - L(inf) | 0.3 | assumed, in line with other stocks |
| Natural mortality - M | 22 | ICES 2006 (Lordan and Gerritsen) |
| Size at maturity (L50) | 0.100 | based on FU15 and FU16 |
| Mature Growth | 50 | based on FU15 |
| Growth - K | 0.2 | assumed, in line with other stocks |
| Growth - L(inf) | 0.000684 | based on Scottish data (Pope and Thomas, 1955) |
| Natural mortality - M | 2.963 | $"$ |
| Length/weight - a |  |  |
| Length/weight - b |  |  |



Figure A.2.1. Effort, catch and catch per unit of effort for Nephrops, Irish otter trawlers in 2005. The boxed and zoomed in plots show a zoomed in view of landings and lpue from the fishery on the Aran Ground.


Figure A.2.2. Time-series of the number of Irish vessels reporting landings of Neprhrops from FU17. The vessels with annual landings $>10 \mathrm{t} / \mathrm{yr}$ can be considered the main participants in the fishery these general account for $\sim 85 \%$ of the total landings.


Figure A.2.3. The time-series of length distributions of Irish vessels landing $\mathbf{> 1 0} \mathbf{t}$ of Nephrops from FU17.


Figure A.2.4. Box plot of the time-series of vessel power in KW of Irish vessels landing $>10 \mathrm{t}$ of Nephrops from FU17.


Figure A.2.5. Monthly landings of Nephrops from FU17 from 1995-2007. The inset shows the average pattern for all years.


Figure A.2.6. The upper panel shows the sex ratio in sampled catches 2003-2008 (error bars = 95\% confidence intervals). The low panel shows the female maturity schedule i.e. percentage at each maturity stage by month.


Figure A.3.1. Anomalies in temperature (upper panel) and salinity (lower panel) for the $53^{\circ} \mathrm{N}$ section running through the Aran Grounds (1999-2006).
a)

b)


Figure A.3.1. Contour and post plots of the a) mean size (phi) and classification based on the Friedman and Sanders (1978) scales and b) sorting ( $\sigma_{\mathrm{g}}$ ) of the sediments on the Aran Grounds based on PSA results from samples collected from 2002-2006.


Figure A.3.2. The bathymetry of the Aran grounds.


Figure B.1.1. Nephrops landings and quota for Ireland since the introduction of TACs in 1987.


Figure B.2.1. Female proportions mature-at-length for FU17. The 95\% confidence limits of the proportions mature-at-length are indicated by the vertical bars. The black curve indicates the model and its standard errors are given by the blue lines. The $L_{50}$ is the estimated length at $50 \%$ maturity and its standard error is given between brackets. Blank plots indicate no sampling took place.

## Stock Annex 7.6: FU16, Porcupine Bank

Stock specific documentation of standard assessment procedures used by ICES.

Stock
Working Group
Date
Revised by

FU16, Porcupine Bank
WGCSE 2010
Version 1, 04/05/2010
Jennifer Doyle

## A. General

## A.1. Stock definition

The Functional Unit for assessment includes some parts of the following ICES Divisions VIIb,c,j,k. The exact stock area is shown on the map below includes the following ICES Statistical rectangles: 31-36 D5-D6; 32-35 D7-D8.


## A.2. Fishery

## France

The French fleet fishing Nephrops in FU16 also fishes in Division VIIg-h and was described in detail in the 1999 WGNEPH Report (ICES, 1999a). The French fleet only lands large Nephrops from this FU. Investigation of the landings data by statistical rectangle carried out by WGNEPH in 2002. These indicated that the majority of the French landings between 1999-2000 were from the south of the Porcupine Bank.

## Ireland

The fishery is mainly seasonal taking place mainly between April and July, landings for the remainder of the year are minimal. Most of the Irish vessels are multi-purpose trawlers and are relatively large (between 20 and 35 m in total length). Irish vessels
land both whole prawns and tails depending on markets from this FU and the sizes of the Irish landings are significantly smaller than those for the French and Spanish fleets. The Irish vessels are mainly using twin-rig trawls. Fishing is often weather dependent (particularly for the smaller vessels), with trip duration varying between seven and ten days. Investigation of the landings data by statistical rectangle provided to the WGNEPH in 2002 indicates that the majority of the Irish landings between 1995 and 2001 were from the south central area of the Porcupine Bank.

The recent spatial distribution of the fishery is shown in Figure 1.

## Spain

The Spanish fishery in the Porcupine area is a typical multi-species fishery, targeting different demersal species, amongst which Nephrops. The fleet, which consists of about 35 vessels, is composed of side-trawlers and is part of the so-called ' 300 fleet' in the Adhesion Treaty of Spain to the EEC in 1986. Within the Porcupine fleet, two components can be distinguished: one consisting of vessels fishing with finfish trawls (average engine power 980 hp ), and the other fishing with Nephrops trawls (average engine power 680 hp ). The average duration of their trips is 15 days, of which $10-12$ are actual fishing days. The major landing port is La Coruña.

The target species for the finfish directed fleet are hake, megrim and anglerfish, with Nephrops as a valued bycatch. Vessels fishing with Nephrops trawls are much more directed towards Nephrops (especially in spring and summer), and fish is a bycatch. These two fleets not are currently disaggregated in the time-series.

## A.3. Ecosystem aspects

Productivity of deep-water Nephrops stocks is generally lower that those on the shelf although individual Nephrops grow to relatively large sizes.

A persistent Taylor column circulation around Porcupine Bank provides an important mechanism for the retention of pelagic eggs and larvae of the various marine species spawning in the area. (Mohn, et al., 2002). The Nephrops stock on the Porcupine Bank is distributed on mud patches in relatively deep waters $200-600 \mathrm{~m}$. It is not know how larvae are retained over these grounds but the Taylor column may help with larval retention.


Figure 1. The spatial distribution of lpue of Nephrops caught by Irish otter trawlers between 20052008 derived using integrated VMS and logbook records.

## B. Data

## B.1. Commercial catch

Commercial catch and effort data is supplied by Ireland, France, Spain and the UK.
These are the countries exploiting the stock.

## B.2. Biological

| BIOLOGICAL PARAMETERS |  |  |
| :--- | :---: | :--- |
| Parameter | Value | Source |
| Discard Survival |  | Discards considered negligible |
| MALES | 0.140 | based on values in other areas (Anon., 1991) |
| Growth - K | 75 | based on maximum sizes observed in samples |
| Growth - L(inf) | 0.2 | Anon, 1990 (estimated) |
| Natural mortality - M | 0.00009 | based on Celtic Sea (FU20-22) |
| Length/weight - a | 3.550 |  |
| Length/weight - b |  |  |
| FEMALES | 0.140 | Not applicable |
| Immature Growth | 75 |  |
| Growth - K | 0.2 |  |
| Growth - L(inf) | 26.2 | Fariña and González Herraiz (2001) |
| Natural mortality - M |  |  |
| Size at maturity | 0.160 | Anon, 1991 |
| Mature Growth | 60 | based on maximum sizes observed in samples |
| Growth - K | 0.2 | As for males |
| Growth - L(inf) | 0.00009 |  |
| Natural mortality - M | 3.550 |  |
| Length/weight - a |  |  |
| Length/weight - b |  |  |

## B.3. Surveys

The only fishery-independent source of data is the Spanish Porcupine trawl survey which commenced in 2001. Further information on this survey is provided in the IBTS Report (ICES, 2010) and in previous IBTS reports. Figure 2 and 3 give gear parameters and spatial distributions of Nephrops catches on the Spanish Porcupine survey.


Figure 2. Door spread, vertical opening and time to settle on the ground between 2004 and 2008.


Figure 3. Distribution of Nephrops norvegicus catches in biomass in Porcupine surveys between 2001 and 2009.

## B.4. Commercial Ipue

The Nephrops fishery on the Porcupine Bank is both seasonal and opportunistic with increased targeting during periods of high Nephrops emergence and good weather.

Effort and lpue data are not standardised, and hence do not take into account vessel capabilities, efficiency, seasonality or other factors that may bias perception of lpue abundance trend over the longer term. The available effort time-series are summarized below:

| Country | First year of <br> effort data | Units | Comment |
| :--- | :---: | :--- | :--- |

Only commercial landings data is available for all countries involved in the fishery.

## B.5. Other relevant data

## C. Historical stock development

An experimental age structured assessment for this stock was carried out by the Nephrops WG in 1993 (ICES, 1993), in 2003 (ICES, 2003) and by the WGHMM (ICES, 2005) in all cases the assessments being considered inadequate. This conclusion was based on poor quality, and unexplainable inconsistencies in the input data. Unknown growth rates and concern about the utility of age based assessment models impeded progress to an accepted assessment. In additional the lack of a time-series of reliable standardised cpue data was also perceived as a problem. This problem has been solved with the developing Porcupine trawl survey-series.

Model used: XSA, LCA
Software used: $\mathrm{n} / \mathrm{r}$
Model Options chosen: No Final model was accepted

## G. Biological reference points

No reference points have been proposed or used for this stock.

## H. Other issues

None.

## I. References

Gerritsen, H. 2009. Working Document 1 ICES Working Group for the Celtic Seas Ecoregion 13-19 May 2009.

## Stock Annex 7.7: Nephrops FU 20-22 (Celtic Sea; VIIfgh)

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Nephrops (Nephrops norvegicus) : Division VIIfgh |
| :--- | :--- |
| Working Group | WGCSE (Working Group for Celtic Seas Ecoregion) |
| Date created | June 2007 |
| Last updated | May 2009 |

## A. General

## A.1. Stock definition

The management area for this stock is delimited in Area VIIfgh (FU 20-22; Figure 1). The management unit is pertinent because of the sedentary feature of Nephrops. However, the sources of recruits are much more poorly defined. There is no evidence that the whole exploited area belongs to the same stock or that there are several patches linked in meta-population sense.

## A.2. Fishery

Nephrops present particular ground features and in the FU 20-22 are known to occur in several areas of muddy sediment and the stock structure is uncertain. The Nephrops fisheries target different areas and have very different size structures in Nephrops catches and landings. These fisheries also have differences in non-Nephrops bycatch composition.

As for all crustaceans, Nephrops grow by successive moults which are to a large extent tied to reproduction. For this species moult occurs twice a year, in spring and autumn until sexual maturity. Once males are sexually mature, they continue to moult twice a year while females moult only once a year in the latter spring/summer right after the hatching of their eggs. In previous references (1970-1980s), it is pointed out that maturation of females happens at a median size of 31 mm CL ( 10 cm of total length) which corresponds to 3.5 years old individuals. There is no specific reference for the sexual maturation of males in the FU 20-22, but biological references on close areas with similar hydrological conditions (FU 15; Western Irish Sea) indicate a first size of functional maturity of 29-31 mm CL.

As reported by the WGNEPH 2004 and the WGSSDS 2005 and 2006, Nephrops in FU $20-22$ is mainly exploited by trawlers from France, Republic of Ireland and UK although the contribution of other countries is lower. The spatial distribution of landings by statistical rectangles are provided below (Figure 2-5). It indicates heterogeneous spatial behaviour of the main fleets.

## France

No major changes have taken place in the fishery for more than fifteen years apart from the implementation of a new mesh regulation in 2000 which increased the minimum codend mesh size from 80 to 100 mm (in fact, the regulation involves to 90 mm mesh size, but 100 mm meshes are adopted aiming to avoid problems with bycatch composition). The 100 mm mesh size also allows them to switch to finfish (cod, whiting, haddock) when Nephrops catch rates are low (e.g. because of diurnal and seasonal variations of catchability for this species or during periods of bad weather). The MLS
applied by the French Producers' Organisations is fixed at 11.5 cm total length (i.e. 35 mm CL). The total number of vessels from the harbours of the South Brittany remains stable (more than 90 declared Nephrops catches from the Celtic Sea in recent years, but around 70 are actually targeting this species). A part of these units (15-20) switch to other Nephrops stocks (FU 16; Porcupine bank; Figure 1) mainly in 2nd and 3rd quarters when the meteorological conditions are favourable. At the opposite, many trawlers (20-30) move towards the FU19 Nephrops (SE and SW Irish coast) mainly in autumn and winter according to difficulties due to weather.

Analytical investigations were carried out on the data collected in 2006 and 2007 involving in the French trawlers. Global indices for fishing effort and lpue provided by this fleet ( 97 trawlers composed by 73 exclusive in Celtic Sea, 15 switching to Porcupine Bank i.e. FU16 and eight also targeting Nephrops in the Bay of Biscay i.e. FU2324) seem to be pertinent: $99 \%$ of vessels* ${ }^{*}$ onths registered for sales at auction can also be found in logbooks ( $94 \%$ of French landings in 2007). In 2006, almost $50 \%$ of French landings occurred in two ICES rectangles (29E2, 30E2; the rectangle 30E2 during the 2 nd quarter concentrated $21 \%$ of yearly landings). In 2007, the contribution of the two rectangles 29E1 and 30E2 was $41 \%$ of yearly landings. In 2008, the rectangles 28 E 1 and 30E2 were represented by $44 \%$ of yearly landings. The peak of production is observed during the 2nd quarter of the year (Figure 4): in 2006, the maximum landings are obtained in June whereas a shift occurred in 2007 (maximum value in May which may be caused by bad meteorological conditions in June). In 2008, the shape of French landings vs. month was bi-modal (May and July were the mostly represented months).

The historical review of French landings shows that the contribution of the rectangle 31 E 3 (concentrating the major part of Irish landings) declined over the last 10 years: from $41 \%$ of total French landings registered in 1999 this contribution is currently less than $10 \%$ (Figure 3). During the last 10 years, the most productive rectangle for French trawlers was 30E2 mainly during the late 2000s: the average annual contribution of this rectangle was around $15 \%$ in the early 2000s, but this proportion reached more than $30 \%$ during the recent years. It seems that the French fleet moved gradually from 31E3 to 30E2 under the steeply increasing concentration of Irish trawlers on the "traditional" Nephrops grounds (Smalls, Labadie).

## Republic of Ireland

More than 60 Irish vessels target Nephrops in the Celtic Sea. In 2007, 95 Irish trawlers were registered as landing Nephrops, but 63 of them exceeded threshold of 10 t (Figure 6). In 2008, 99 Irish vessels reported landings from this area whereas 67 of them landed more than 10 t . The fishery presents a more typical seasonal profile than the French vessels and most of the landings are made between March and July. These vessels are mid-size multi-purpose trawlers, with a length of $18-23 \mathrm{~m}$ and engine power between 250 and 350 kW . Many of the vessels switch between FU15 and FU20-22, depending on the tides in the Irish Sea. Other vessels switch from targeting finfish in the winter to Nephrops in the spring and early summer. The mesh size used by Irish vessels is 80 mm , and increasingly these vessels are using twin trawls. The MLS applied by Irish trawlers is the European one fixed at 8.5 cm total length (i.e. 25 mm CL ).

The Irish landings seem to be more concentrated spatially than the French. During the period 2003-2006, 63-67\% of the Irish nominal landings were provided by one ICES rectangle (31E3). The Irish fishing effort is located more northerly than the French one.

## UK

The UK fishery in the Celtic Sea has generally remained unchanged. Since the early 2000s, the number of UK Nephrops directed vessels has increased from around 10 to 15 , but their contributions in total landings remains minor (usually less than 50 t of landings). The maximum historical value of UK landings is reported in 2008 (242 t).

## A.3. Ecosystem aspects

Nephrops occur in discrete patches where the sediment is suitable for them to construct their burrows. There is a larval phase of long duration where there may be some mixing with Nephrops from other areas depending on the oceanographic conditions, but the mechanisms for this in the Celtic Sea are not currently known.

Cod has been identified as a predator of Nephrops in some areas, and the generally low level of the cod stock is likely to have resulted in reduced predation on Nephrops.

## B. Data

## B.1. Commercial catch

Landings are reported mainly by France and the Republic of Ireland. French landings fluctuated between 2000 and 3800 t . Irish landings rose from around 500 to more than 2000 t in the last 15 years. The highest value of Irish landings is observed in 2007 (more than 3200 t ). A part of this trend is due to greater accuracy of reporting mainly after the end of the late 1990s. The contribution of French landings has gradually decreased from $80-90 \%$ at the end of 1980 s to $50-60 \%$ at the beginning of 2000 s . Between 2004 and 2005, French landings remained stable whilst Irish landings steeply increased and the total harvested quantity was the highest during the last decade. For the first time, in 2007, the Irish ladings exceeded the French ones ( 3230 t against 2080 t). This may be caused by constraints linked to the international context affecting fuel prices for fishing vessels. The overall fishing profile remains typically seasonal with a dominance of the 2 nd and 3rd quarters ( $60-70 \%$; the other quarters are less productive because of meteorological conditions and of less accessibility of females due to burrowing).

During the recent years, the evolution of the French fishing effort and lpue was sometimes considerably different from the evolution of the same indicators for the Irish fleet (e.g. between 2004 and 2005: $-5 \%$ of fishing effort and $+2 \%$ of lpue for French trawlers against $+50 \%$ of fishing effort and $+25 \%$ of lpue for Irish trawlers). In 2007, an increase occurred for lpue values of both main fleets: a slight upwards trend of French trawlers ( $+13 \%$ associated to a strong reduction of the fishing effort: $-25 \%$ whereas the total number of vessels remained almost stable) and a steep one for the Irish fleet ( $+36 \%$ coinciding with $+31 \%$ of the fishing effort which was displayed by an increasing number of trawlers operating in the Celtic Sea: $+19 \%$ between 2006 and 2007). This underlines the divergence of features of the targeting vessels for each country and indicates the great heterogeneity of the area. A direct comparison between both countries cannot be undertaken because the fishing effort is not available in the same unit (France: otter trawlers getting at least $10 \%$ of their total landings by targeting this species; Ireland: otter trawl vessels where $>30 \%$ of monthly landings in live weight were Nephrops). Furthermore, the actual fishing areas are different and the Irish fleet is more restricted spatially as already reported by WGSSDS 2005-2008.

## B.2. Biological

## Natural mortality and maturity-at-age

A natural mortality of 0.3 is applied to all Nephrops males whereas the mortality of females changes at the size of first maturity (occurring at 31 mm CL as explained previously): a value of 0.2 is usually applied on mature individuals.

The L2AGE slicing program usually applied on Nephrops stocks allocates length classes into age groups by assuming von Bertalanffy model of individual growth. This slicing is applied to length distributions by sex. All parameters, $L \infty$ and $K$ by sex, calculated mean sizes by age for each sex, natural mortality and maturity by sex (assumed to be knife-edged for males and s-shaped for females) and combined are given below.

Table 1. Nephrops FU20-22 (Celtic Sea). Individual growth, natural mortality, maturity parameters by sex.

| age |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Size | males | 11 | 20 | 27 | 34 | 39 | 44 | 47 | 51 |
| (CL mm) mm | females | 11 | 20 | 27 | 32 | 33 | 35 | 36 | 37 |
| M | males | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 |
|  | females | 0.3 | 0.3 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
|  | combined | 0.3 | 0.3 | 0.3 | 0.25 | 0.25 | 0.25 | 0.25 | 0.25 |
| Maturity | males | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 |
|  | females | 0 | 0 | 0 | 0.5 | 1 | 1 | 1 | 1 |
|  | combined | 0 | 0 | 0.5 | 0.75 | 1 | 1 | 1 | 1 |

## Biological sampling

Landings: The total French landings have been available since 1983 (on quarterly basis since 1987) whereas the Irish series began in 1987 (on quarterly basis since 1995).

Lpue and fishing effort: Lpue series are provided since 1987 in France whilst Irish data are available over 1996. It has to be noted that the French and Irish method of calculation of the fishing effort are not carried out by the same way (threshold of $10 \%$ in weight for Nephrops on total landings applied for French trawlers whereas 30\% is the threshold used for Irish fleet), thus a direct comparison of those indices is not appropriate.

DLF of landings: French sampling plan at auction started in 1983, but only after 1986 the data can be used on quarterly basis. The Irish plan as written previously began in 2002 (in fact, solely 2003 has been entirely sampled in the FU20-22 area; 2002 data involving the whole Management Area M: see processing by WGSSDS 2006; two quarters were not sampled in 2004 and 2005: see processing by WGSSDS 2006). For French landings, the increasing proportion of tailed individuals (see below) and the inappropriate method of sampling before the end of 2007 provided

DLF of discards: French estimation of discards occurred only in three separate years (1985, 1991 and 1997), but only the data collected in 1997 can be included in analytical investigations. The available dataset is given for only one year of discard sampling (1997) because of unavailable quarterly data for landings for the first year of discard
sampling (1985) whereas data collected in 1991 were considered as unreliable (samples sorted by fishermen). Irish sampling has been undertaken since 2002 (lack of information for two quarters in 2004; see processing by WGSSDS 2006).

Length compositions of the landings by sex are provided for the two main fleets, but the time-series are different. Sampling of French landings since 1984 has provided length frequencies by sex on a monthly basis. Due to uncertainty of the older datasets, the data for 1984-1986 were omitted from further analysis. The Irish sampling program was launched in 2002 under the EU DCR and gave length frequencies for the period 2002-2006 (after simulation undertaken for some missing information in 2004 as explained during WGSSDS 2006).

French estimation of discards occurred only in several separate years (1985, 1991 and 1997; in 2005, samples for two quarters, 3rd and 4th, were also provided), but only the data collected in 1997 can be included in analytical investigations because of unavailable quarterly data on landings for the first year of discard sampling (1985) whereas data collected in 1991 were considered as unreliable (samples sorted by fishermen not representative of the discarding behaviour of the whole fleet). The 1997 French plan onboard showed high spatial and temporal variability of discard sizecomposition vs. that of landings (CV>30\%). The Irish sampling launched under DCR gave results as presented by Table 2.

The heterogeneity of the dataset in addition to that of the harvested area by each country affects the discard rate by fleet: it was higher for French vessels: 65\% in 1997 against $37 \%$ for Irish in 2003 (the only one year with sampling, but only $11 \%$ during the quarters 2 and 3 in 2004) and by sex (stronger in the case of females growing less quickly).

Table 2. FU 20-22 Irish Sampling Summary.

| Year | Quarter | Number of samples |  |  | Numbers Measured |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Discards | Landings | Catch | Discards | Landings |
| 2003 | 1 | 1 | 1 |  | 186 | 417 |  |
|  | 2 | 5 | 5 |  | 4057 | 3016 |  |
|  | 3 | 3 | 3 |  | 2535 | 3638 |  |
|  | 4 | 2 | 1 |  | 996 | 528 |  |
| 2004 | 1 | 0 | 0 |  | 0 | 0 |  |
|  | 2 | 3 | 2 |  | 1634 | 2781 |  |
|  | 3 | 7 | 6 |  | 4284 | 7171 |  |
|  | 4 | 0 | 0 |  | 0 | 0 |  |
| 2005 | 1 | 1 | 1 |  | 1330 | 2271 |  |
|  | 2 | 2 | 2 |  | 2208 | 3238 |  |
|  | 3 | 2 | 0 |  | 1634 | 0 |  |
|  | 4 | 2 | 0 |  | 1627 | 0 |  |
| 2006 | 1 | 2 | 1 | 2 | 1891 | 1152 | 2252 |
|  | 2 | 10 | 2 | 2 | 7241 | 1049 | 363 |
|  | 3 | 5 | 1 | 0 | 3178 | 1101 | 0 |
|  | 4 | 9 | 0 | 0 | 8266 | 0 | 0 |
| 2007 | 1 | 1 | 3 | 0 | 767 | 770 | 0 |
|  | 2 | 12 | 0 | 0 | 9648 | 0 | 0 |
|  | 3 | 15 | 4 | 2 | 7784 | 1862 | 411 |
|  | 4 | 6 | 5 | 0 | 1959 | 1417 | 0 |
| 2008 | 1 | 2 | 5 |  | 680 | 1758 |  |
|  | 2 | 10 | 13 |  | 3409 | 5333 |  |
|  | 3 | 3 | 2 |  | 878 | 546 |  |
|  | 4 | 4 | 4 |  | 1356 | 1573 |  |

## Extrapolations

## Landings: DLF of tailed Nephrops

The WGCSE 2009 pointed out a significantly increasing proportion of tailed individuals in French landings whereas this proportion was already high for Irish trawlers. In 2008, $20 \%$ of total French landings involved in tailed Nephrops ( $19 \%$ in 2007, $15 \%$ in 2006 and $11 \%$ in 2005 ; less than $5 \%$ until the beginning of 2000s). The overall upwards trend is illustrated by the Figure 7 presenting also monthly tailed fractions (after conversion of weight of tails to total one).

The seasonal variability of tailed Nephrops may be explained by biological features of the species (two peaks appear by year corresponding to the two moulting periods, spring and winter) and by the particular conditions of trips (12-15 days) compromising the conservation of Nephrops. As regards to the annual increasing proportion of tails ( $96 \%$ explained by using an exponential function), industry explained it by the economic difficulties of the vessels because of the rapidly increasing fuel prices. Tailed individuals are intended to compensate this loss for the crew participation at the total investment by trip. As the European MLS for FU20-22 Nephrops is fixed at 8.5 cm of total length ( 25 mm CL) and the MLS retained by the French Producers' Or-
ganizations is equal to $11.5 \mathrm{~cm}(35 \mathrm{~mm} C L)$, it was expected that tailed individuals should be comprised between these two sizes.

Before the end of 2007, the tailed Nephrops could not be sampled at auction and, as the sampling onboard remains difficult to apply routinely (long trip duration for French trawlers), the problem was partially tackled by apportioning tailed individuals to the smallest category of landings at auction. Since the end of 2007, new biometric relationships established during the EVHOE survey have been used: they allow fitting CL vs. 2nd abdominal segment of tail by sex (Figure 8). The DLF of French landings for 2008 were estimated by two ways: one using the extrapolations from tails to CL, the other apportioning tails to the small category as for previous years. The resulting difference appears relevant (Figure 9): in 2008, 46 million Nephrops were provided by the previous method whereas 58 million were estimated by including tails ( $+28 \%$ ). Almost $30 \%$ of landed individuals were below the French Producers' Organization MLS, but no Nephrops was undersized compared with European MLS. Moreover, the sex ratio seems to be affected by the tailing practice: $13 \%$ of Nephrops ( 7.4 million) were females although this percentage would be $7 \%$ ( 3.2 million) under the previous method. The mean size of French landings for 2008 decreases at around $2.5-5 \mathrm{~mm}$ CL by sex when tails are involved by sampling. However, the mean CL for 2008 remains larger than the Irish one.

Table 3. Nephrops in VIIfgh. Mean sizes (CL in mm) of French and Irish landings for 2008. French values are calculated (1) including the samples involving in tailed individuals and (2) using the previous method (no sampling of tails; the total tailed proportion was apportioned in the smallest category of entire Nephrops at auction).

| French sampling |  | Irish sampling |  |  |  |  |
| :--- | :---: | :--- | :--- | :--- | :---: | :---: |
| Males |  | Females | Total | Males | Females | Total |
| 37.6 | 34.7 | 37.2 |  | 32.0 | 29.7 | 31.1 |
| 40.1 | 39.6 | 40.1 |  |  |  |  |

This result emphasizes the WGSSDS 2008 conclusion that the size composition may be overestimated when raised to the composition of entire individuals.

Discards: years with no sampling onboard

## Generalities

As the sampling plan for both countries was not routinely undertaken, the whole time-series of landings by quarter either for the French fleet (years 1987-2007) or for the Irish one (years 1995-2007, years 1987-1994 are only represented by annual landings) misses information. Therefore, a methodology of extrapolation from sampled data to years or quarters with no information was developed (see WD 1; WGSSDS 2007).

The main concepts of the derivation (back-calculation) are summarized as:
1 ) The first step involves applying hand-sorting selection of retained catches which is explained by s-shaped (logistic) function vs. size. As statistically tested by fleet, the hand-sorting function is stable within-quarter for given parameters of the exploitation pattern (if mesh size and MLS remain constant within period).
2 ) The second step consists in removing undersized individuals unusual in landings which can generate unreliably extreme values of discards due to
sampling problems (very high CV of landings for the extreme size classes). Hence, size classes less than a tested threshold (e.g. 1 or $5 \%$ of cumulative landings) were eliminated.
3 ) The third step allows the generation of missing size classes by applying a probability density function which can be symmetrical or not. The whole calculation is based on multiple maximum likelihood function according to the number of missing years. Relationship as between mean sizes of landings and of discards tested on the FU23-24 Nephrops (Bay of Biscay; WGHMM) can also be included in the final fitting.

## Particularities for FU20-22 Nephrops stock

The approach summarized above was already developed on the FU23-24 Nephrops stock (Bay of Biscay) and its validation was investigated during the WGHMM 2007 (Figure 10-14). The WGSSDS 2007 examined statistical formulation and validation of this method on French (years 1987-2006) and Irish (years 2002-2006, investigation by quarter) discards for FU20-22. There are some differences from the calculation applied on the Bay of Biscay as:

1 ) The available French dataset is given for only one year of discard sampling (1997). It means that the hand-sorting s-shaped curves by quarter are calculated on only one year ${ }^{1}$ instead of six in the case of the Bay of Biscay stock.
2 ) The cumulative percentage level for removing of undersized generated discards (see above: 2nd stage) is fixed at $5 \%$ for French data and $1 \%$ for Irish data (also $1 \%$ for the Bay of Biscay Nephrops stock). In the case of the French fishery in Celtic Sea, this can be justified by the high variability of landing samples between trips (higher coefficients of variation at auction because of higher heterogeneity of the fished area and of long duration of trips i.e. 12-15 days and, hence, less availability of samples at auction).

3 ) For the French discards, with only one year of discard sampling, the initial value of the parameter Lm cannot be assumed to be equal to any expected mean size of discards $v s$. mean size of landings (see above 3rd stage). Furthermore, the interval in which Lm should be contained is not statistically calculable. Hence, Lm is initially introduced as the size corresponding to the maximum number of discarded individuals as provided by the 2nd stage of calculation (i.e. after removing extremely high values of discards obtained after the 1st stage: hand-sorting logistic function). Its interval is built by using an a priori coefficient of variation around the initial Lm (CV of 0.10 and 0.20 were tested). For the Irish data, no constraint on relationship between mean sizes of discards and landings was set because of lack of any information on that due to the short time-series.

4 ) (4) The large mesh size of the French vessels in the FU20-22 area indicates that the distribution of length frequencies of discards is probably no symmetrical because of selectivity effects which should be more significant than for the FU23-24 stock or for the Irish trawlers in the FU20-22.
5 ) For French discards, the absence of reference about any relationship between mean sizes of landings and discards at the opposite of the Bay of

[^21]Biscay, implies that the final fitting aims to provide the more linear as possible relationship (after log-log transformation) with only one reference point (year 1997). Hence, the optimisation is more based on geometric concept than on statistical one.

## 1st stage: the s-shaped hand-sorting curve

Let j be a year with no dataset on discards. By quarter $k$, the number of discarded individuals by sex ( $m$ or $f$ ) and by size $L, \mathrm{ND}_{\mathrm{jklm}}$ (or $\mathrm{ND}_{\mathrm{jklf}}$ ), is not calculated on data provided from other years, but from the number of landed individuals NLiklm (or $N L i k l f)$ during the same year, quarter $k$, sex ( $m$ or $f$ ) and size $L$ :

$$
\begin{align*}
& N D_{j k l m}=N L_{j k l m} \cdot \exp \left(-\alpha_{k} \cdot\left(L-L 50_{k}\right)\right) \text { or } \\
& N D_{j k l f}=N L_{j k l f \cdot} \cdot \exp \left(-\alpha_{k .}\left(L-L 50_{k}\right)\right) \tag{1}
\end{align*}
$$

$\alpha_{\mathrm{k}}$ and L50k ${ }_{\mathrm{k}}$ are the parameters of the s-shaped curve (logistic model) fitted by quarter k describing the commercial Nephrops hand-sorting onboard. For this fitting, both sexes are combined and the dependent variable is expressed by the number of landed individuals for size $L$ and the independent one is the total number of catches by size L for the years with discard sampling onboard.

The estimates $\alpha_{\mathrm{k}}$ and L50k were calculated by assuming the stability of hand-sorting process onboard if mesh size and MLS remain unchanged. The short Irish time-series 2002-2006 was considered as a common dataset, but, for the French trawlers, the overall time-series was divided into three periods:

1 ) Years 1987-1990: The results of sampling carried out in 1985 are not available on computing support. Thus, there is no formal information if the hand-sorting onboard could be approximated by the more recent parameters of 1990s. $\alpha$ and L50 were not got fixed, but their values were estimated by the multiple likelihood function as for the parameters of the probability density by year (see below).
2 ) Years 1991-1999: The hand-sorting was fitted on data from 1997 (1991 data were not representative of the whole fleet). The missing data of years 19911996 and 1998-1999 were therefore estimated.
3 ) Years 2000-2006: Because of the mesh size change, the hand-sorting should be different from 1997 sampling data. However, there is no new information for the 1st and 2nd quarters (the 2005 sampling plan provided relevant results only for the 3rd and 4th quarters). Hence, $\alpha$ and L50 for the first two quarters were fixed equal to 1997 parameters, but the simulation for the other two quarters is based on 2005 data.

## 2nd stage: removing of unreliable size classes of discards

This derivation approach reduces interdependence between yearly datasets which may induce lack of contrast in recruitment time-series. In spite of that, some inconveniencies of the new approach have to be taken into account: (1) the hand-sorting onboard s-shaped curve implies that, for a given size class, no calculation of discards is possible while there is no landed individuals and (2) the exponential expression gives extremely unreliable high values of discards when undersized individuals are sampled in landings (mainly because of hand-sorting deviation due to sampling rate not representative for extreme size classes).

1) Undersized individuals unusual in landings. As written previously, undersized Nephrops sampled in landings should produce unreliable high discarded amounts by size because of the exponential calculation. All size classes representing less than a minimum cumulative percentage level in landings by year were removed ( $5 \%$ for French landings, $1 \%$ for Irish landings).
2 ) Discarded individuals by size exceeding observed mean ratios discards/landings. Generated discarded numbers were removed when the calculated ratio discards/landings by size (decreasing function vs. size) exceeded observed mean ratios by size ${ }^{2}$. Almost all size classes involved by (2) were already removed by (1). This operation was added at the aim of elimination of not normally high ratios discards/landings for large sizes (which has a little impact on total discarded number due to the s-shaped function of handsorting).

This calculation process retains only a part of the initial hand-sorting generated distributions of discards mainly the decreasing part of discarded individuals.

## 3rd stage: simulation of densities of probability of discarded individuals (yearly distribution for French and quarterly for Irish discards)

Finally, the assumed distribution of discards for the whole range of sizes was calculated from the descending part. This process needs to input the probability density of discards given by:

$$
\begin{equation*}
\varphi(L)=\frac{\alpha}{1+\exp (\beta \cdot(L-L m)} \tag{2}
\end{equation*}
$$

where $\alpha, \beta, \mathrm{Lm}$ are coefficients of the distribution $(\phi(\mathrm{L})=\alpha / 2$ when $\mathrm{L}=\mathrm{Lm})$.
Because of the assumed skewness for the French discard distribution, as explained above, the whole function of the probability density is approximated by:

$$
\begin{align*}
& \varphi(L)=\frac{\alpha}{1+\exp (-\gamma \beta \cdot(L-L m)} \text { for } L \leq L m \\
& \varphi(L)=\frac{\alpha}{1+\exp (\beta .(L-L m)} \text { for } L>\operatorname{Lm} \tag{3}
\end{align*}
$$

with a complementary coefficient $\gamma$ : if $\gamma=1$ the whole probability density is symmetrical, if $\gamma<1$ the skewness of the distribution is positive if $\gamma>1$ the skewness is negative ( $\gamma=1$ for Irish discards, $\gamma \neq 1$ for French discards).

The fitting of $\phi(\mathrm{L})$ is processed on two stages:

- Lm and $\alpha$ are fixed: $\alpha$ is initially fixed at $2^{*} \phi$ max which is the maximum frequency retained after the 2nd stage of calculation (see above), Lm is fixed at the size corresponding to the maximum number of discarded individuals as provided by the 2nd stage of calculation (see previously) and, hence, $\beta$ is given by:

$$
\begin{equation*}
\beta=\frac{1}{n} \sum_{L=L \min }^{L \min +n-1} \ln \left[2 \cdot \frac{\varphi \max }{\varphi(L)}-1\right]^{\frac{1}{L-L m}} \tag{4}
\end{equation*}
$$

[^22](Lmin= first size represented by not null individuals and $\mathrm{n}=$ number of total size classes with discards different from zero).
All parameters are estimated: $\alpha, \beta, \mathrm{Lm}$ got obtained by the 1 st stage are input for the final calculation using Newton cancellation of gradient and assuming stochastic approach for Lm. Lm is assumed to be included in the interval defined accordingly to an a priori CV of Lm (see above) ${ }^{3}$.

Otherwise, the final run includes constraints as:

- The sum of frequencies for descending part of distribution is equal to that calculated by the model i.e. the retained values of the 2nd stage of calculation described previously are assumed to be reliable.
- $\mathrm{Lm} \geq \mathrm{Lmin} \quad\left[\mathrm{Lmin}=(1-\mathrm{Z} 1-\alpha / 2 . \mathrm{CV})^{*} \mathrm{Lm}\right] \quad$ (usually: $\alpha=0.05=>\mathrm{Z}_{1}$ $\alpha / 2=1.96$ )
- $\mathrm{Lm} \leq \mathrm{Lmax} \quad\left[\mathrm{Lmax}=\left(1+\mathrm{Z}_{1-\alpha / 2 . \mathrm{CV}}\right)^{*} \mathrm{Lm}\right]$
- For French discards, the coefficient of determination of the relationship between the mean sizes of landings and the mean sizes of discards for missing years has to be as close as possible to 1 (with no possibility of statistical test because of only one year dataset).


## Statistical formulation and validation

## Calculation of variances

## Matrix of variances-covariances of model parameters

The Generalized Reduced Gradient and the Complex method do not give an estimate of the matrix of variances-covariances of the four (three for Irish) parameters. In this case, it is usually recommended to apply non-parametric techniques such as the Bootstrap method. The calculation can also be carried out according to parametric procedure (Lin, 1987; Fifas and Berthou, 1999; Fifas et al., 2004) using Jacobian matrix (i.e. matrix of partial derivatives of the objective).

The matrix of variances-covariances is obtained by the following relationship:
$[\mathrm{M}]=\mathrm{s}^{2} .[\mathrm{I}]^{-1}$
with:
$[\mathrm{M}]=$ matrix of variances-covariances; $[\mathrm{I}]^{-1}=$ inverse of matrix of information; $\mathrm{s}^{2}=\mathrm{sum}$ of mean residual squares of the fitted function $\left(s^{2}=\right.$ SCE/DDL $\left.{ }^{4}\right)$ :

$$
\begin{equation*}
S C E=-\sum_{i=1}^{L j<L m}\left[\varphi\left(L_{i}\right)-\frac{\alpha}{1+\exp \left(-\gamma \beta \cdot\left(L_{i}-L m\right)\right.}\right]^{2}+\sum_{i=j+1}^{L j>=L m}\left[\varphi\left(L_{i}\right)-\frac{\alpha}{1+\exp \left(\beta \cdot\left(L_{i}-L m\right)\right.}\right]^{2} \tag{6}
\end{equation*}
$$

The matrix of information is obtained by:

$$
\begin{equation*}
[\mathrm{I}]=[\mathrm{J}]^{\prime} \cdot[\mathrm{J}] \tag{7}
\end{equation*}
$$

[^23][J] is the Jacobian matrix (nc rows and 4 columns for French data, 3 for Irish):
\[

[J]\left[$$
\begin{array}{l}
\frac{\partial \varphi\left(L_{1}\right)}{\partial \alpha} \frac{\partial \varphi\left(L_{1}\right)}{\partial \beta} \frac{\partial \varphi\left(L_{1}\right)}{\partial \gamma} \frac{\partial \varphi\left(L_{1}\right)}{\partial L m}  \tag{8}\\
\frac{\partial \varphi\left(L_{2}\right)}{\partial \alpha} \frac{\partial \varphi\left(L_{2}\right)}{\partial \beta} \frac{\partial \varphi\left(L_{2}\right)}{\partial \gamma} \frac{\partial \varphi\left(L_{2}\right)}{\partial L m} \\
\frac{\partial \dot{\varphi}\left(L_{n c}\right)}{\partial \alpha} \frac{\partial \dot{\varphi}\left(L_{n c}\right)}{\partial \beta} \dot{\partial \dot{\varphi}\left(L_{n c}\right)} \frac{\partial \varphi\left(L_{n c}\right)}{\partial \gamma}
\end{array}
$$\right]
\]

[J]' is the transpose of [J], the partial derivatives of the equation [8], also defined as absolute coefficients of sensitivity of order 1 written as $a(\alpha), a(\beta), a(\gamma), a(L m)$ are given below:

$$
\begin{align*}
& \frac{\partial \varphi(L)}{\partial \alpha}=\frac{\varphi(L)}{\alpha}  \tag{9}\\
& \frac{\partial \varphi(L)}{\partial \beta}=\gamma \cdot(L-L m) \cdot \varphi(L) \cdot\left(1-\frac{\varphi(L)}{\alpha}\right) \text { if } \mathrm{L} \leq \mathrm{Lm}  \tag{10a}\\
& \frac{\partial \varphi(L)}{\partial \beta}=-(L-L m) \cdot \varphi(L) \cdot\left(1-\frac{\varphi(L)}{\alpha}\right) \text { if } \mathrm{L}>\mathrm{Lm}  \tag{10b}\\
& \frac{\partial \varphi(L)}{\partial \gamma}=\beta \cdot(L-L m) \cdot \varphi(L) \cdot\left(1-\frac{\varphi(L)}{\alpha}\right) \text { if } \mathrm{L} \leq \mathrm{Lm}  \tag{11a}\\
& \frac{\partial \varphi(L)}{\partial \gamma}=0 \text { if } \mathrm{L}>\mathrm{Lm}  \tag{11b}\\
& \frac{\partial \varphi(L)}{\partial L m}=-\beta \cdot \gamma \cdot \varphi(L) \cdot\left(1-\frac{\varphi(L)}{\alpha}\right) \text { if } \mathrm{L} \leq \mathrm{Lm}  \tag{12a}\\
& \frac{\partial \varphi(L)}{\partial \gamma}=\beta \cdot \varphi(L) \cdot\left(1-\frac{\varphi(L)}{\alpha}\right) \text { if } \mathrm{L}>\mathrm{Lm} \tag{12b}
\end{align*}
$$

## Uncertainty of simulated discards

The matrix of variances-covariances of the four (three for Irish) parameters of the model and the use of partial derivatives of order 1 provide an approximate calculation of the variance of the variable $\Psi(\mathrm{L})$ corresponding to simulated discards vs. size L. This procedure is based on limited developments of order 1 in Taylor's series (called Delta methods: Laurec, 1986; Laurec and Mesnil, 1987; Chevaillier, 1990; Chevaillier and Laurec, 1990; Fifas and Berthou, 1999; Fifas et al., 2004).

By using Taylor's polynomial on a function $\Phi$ against parameters $\theta_{1}, \theta_{2}, \ldots, \theta_{\mathrm{k}}$ it is possible to present the variance of $\Phi$ by:

$$
\begin{equation*}
V[\Phi] \approx \sum_{i=1}^{k}\left(\frac{\partial \Phi}{\partial \theta i}\right)^{2} \cdot V\left[\theta_{i}\right]+2 \cdot \sum_{i=1}^{k-1} \sum_{j=i+1}^{k} \frac{\partial \Phi}{\partial \theta i} \cdot \frac{\partial \Phi}{\partial \theta j} \operatorname{Cov}\left[\theta_{i}, \theta_{j}\right] \tag{13}
\end{equation*}
$$

Then, the variance of simulated discards vs. size, $\mathrm{V}[\Psi(\mathrm{L})]$, is written as:

$$
\begin{align*}
& V[\Psi(L)] \approx a(\alpha)^{2} \cdot V[\alpha]+a(\beta)^{2} \cdot V[\beta]+a(\gamma)^{2} \cdot V[\gamma]+a(L m)^{2} \cdot V[L m]+2 \cdot a(\alpha) \cdot a(\beta) \cdot \operatorname{Cov}[\alpha, \beta]+ \\
& \text { 2.a( } \alpha) \cdot a(\gamma) \cdot \operatorname{Cov}[\alpha, \gamma]+2 \cdot a(\alpha) \cdot a(L m) \cdot \operatorname{Cov}[\alpha, L m]+2 \cdot a(\beta) \cdot a(\gamma) \cdot \operatorname{Cov}[\beta, \gamma]+2 \cdot a(\beta) \cdot a(L m) \cdot \operatorname{Cov}[\beta, L m]+ \\
& \text { 2.a(}(\gamma) \cdot a(L m) \cdot \operatorname{Cov}[\gamma, L m] \tag{14}
\end{align*}
$$

where the absolute coefficients of sensitivity of order 1 (partial derivatives) are defined above (equations [9] to [12]).

## Validation

The generated by simulation values are tested against discards estimated by sampling. This procedure is undertaken on French data of 1997 and also on available Irish set (all quarters of 2003, 2004-Q2, 2004-Q3, 2005-Q1, 2005-Q2, 2006 apart from Q4 i.e. 11 quarters). As performed for the Bay of Biscay Nephrops stock, this validation involves in three main stages (Figure 10-14): (1) Examination of the total amount of discards calculated by simulation that should not be significantly different from that obtained by sampling. (2) Test by linear regression performed on simulated numbers vs. size as dependent variable against sampled numbers as independent one. The slope of this relationship should not be significantly different from 1 (bisecting line) and the intercept should not be significantly different from 0. (3) Test of cumulative frequencies of the sets, sampled and simulated, using non parametric approaches such as Kolmogorov-Smirnov.

## Results

## Hand-sorting s-shaped curves

The French and Irish hand-sorting logistic curves estimated by sampling are provided by Figure 15. In the Table 4, are also presented the French parameters involving in years 1987-1990 (simulated by the multiple likelihood function applied for probability density of discards; see above).

Table 4. Summary of parameters of s-shaped hand-sorting curves.

|  | FR (years 1987-1990) |  | FR (year 1997) |  | IRL (years 2003-2005) |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| quarter | $\alpha$ | L50 | $\alpha$ | L50 |  | $\alpha$ |
| Q1 | 0.797 | 32.685 | 1.006 | 32.776 | 0.480 | 25.876 |
| Q2 | 0.494 | 35.573 | 0.718 | 36.019 | 0.426 | 26.016 |
| Q3 | 0.331 | 32.227 | 0.851 | 33.654 | 0.559 | 25.785 |
| Q4 | 0.697 | 31.138 | 0.815 | 32.381 | 0.412 | 24.886 |

These values indicate the high heterogeneity between the two fleets which accentuates the a priori high spatial heterogeneity of the targeted resource. Some weak differences are observed between the simulated values $\alpha$ and L50 of the first French period (1987-1990) and the sampling of 1997. Nevertheless, these parameters are given by deterministic way; therefore, there is no possibility of further statistical comparison.

## Estimates of French discards

Estimates of French discards (1987-2006), total number of discarded individuals, parameters $\alpha, \beta, \gamma$ and Lm and corresponding coefficients of variation (CV, in \%), are given below (Table 5). The Table 6 and Figure 16 present discard rates by sex and combined for the overall time-series.

Table 5. French Nephrops trawlers, Celtic Sea (FU20-22). Estimates of discards, coefficients of model and coefficients of variation of parameters.

| year | disc | $\mathrm{CV}($ disc $)$ | Lm | $\mathrm{CV}(\mathrm{Lm})$ | $\boldsymbol{\alpha}$ | $\mathrm{CV}(\boldsymbol{\alpha})$ | $\boldsymbol{\beta}$ | $\mathrm{CV}(\boldsymbol{\beta})$ | $\boldsymbol{\gamma}$ | $\mathrm{CV}(\boldsymbol{\gamma})$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 125752 | 4.62 | 30.278 | 3.25 | 25773 | 13.79 | 0.293 | 32.11 | 0.768 | 44.61 |
| 1988 | 425396 | 4.88 | 28.917 | 5.28 | 59518 | 16.97 | 0.260 | 39.24 | 0.534 | 56.57 |
| 1989 | 99536 | 4.02 | 31.061 | 4.36 | 14417 | 13.86 | 0.221 | 33.01 | 0.740 | 45.69 |
| 1990 | 81530 | 8.74 | 30.579 | 8.28 | 12219 | 28.86 | 0.221 | 61.77 | 0.866 | 92.51 |
| 1991 | 389726 | 5.69 | 29.479 | 5.70 | 57932 | 18.85 | 0.218 | 40.78 | 0.868 | 60.75 |
| 1992 | 377075 | 18.48 | 30.752 | 14.57 | 61039 | 58.97 | 0.314 | 142.51 | 0.534 | 193.98 |
| 1993 | 118210 | 199.42 | 31.299 | 147.10 | 20679 | 612.24 | 0.258 | 1356.53 | 0.879 | 1956.90 |
| 1994 | 93687 | 7.62 | 31.438 | 6.77 | 14384 | 24.84 | 0.232 | 54.91 | 0.830 | 79.80 |
| 1995 | 131541 | 136.57 | 31.808 | 95.39 | 25096 | 418.52 | 0.273 | 880.20 | 0.808 | 1323.18 |
| 1996 | 82811 | 6.05 | 32.357 | 5.61 | 12121 | 20.20 | 0.255 | 49.20 | 0.637 | 66.91 |
| 1997 | 96612 | 6.21 | 32.403 | 2.11 | 18050 | 15.36 | 0.673 | 46.01 | 0.397 | 55.62 |
| 1998 | 30494 | 7.62 | 31.393 | 10.98 | 3453 | 28.85 | 0.161 | 61.94 | 0.893 | 94.65 |
| 1999 | 36900 | 12.14 | 31.827 | 10.67 | 5618 | 40.01 | 0.236 | 84.90 | 0.791 | 127.28 |
| 2000 | 22234 | 46.41 | 33.790 | 56.24 | 2655 | 171.90 | 0.175 | 359.92 | 0.863 | 552.62 |
| 2001 | 98962 | 5.59 | 31.766 | 7.43 | 11594 | 20.94 | 0.191 | 46.64 | 0.682 | 69.25 |
| 2002 | 34283 | 18.42 | 33.466 | 21.52 | 4223 | 66.86 | 0.193 | 150.64 | 0.762 | 217.87 |
| 2003 | 59692 | 4.73 | 34.452 | 3.48 | 9659 | 15.04 | 0.285 | 36.31 | 0.638 | 49.26 |
| 2004 | 29493 | 9.36 | 33.546 | 9.20 | 4050 | 32.24 | 0.202 | 69.23 | 0.874 | 103.22 |
| 2005 | 15097 | 18.92 | 34.739 | 17.57 | 2098 | 65.03 | 0.205 | 136.51 | 0.873 | 206.98 |
| 2006 | 17286 | 6.86 | 36.327 | 7.29 | 2350 | 24.93 | 0.238 | 64.77 | 0.530 | 85.17 |

Note: the sampled year 1997 is given in bold and italic fonts whereas in coloured fonts are presented the years for which the model based on the probability density seems to be inappropriate (years 1993, 1995, 2000; extremely high CV of parameters and discarded numbers). The total discarded number cited for 1997 is the value obtained by sampling.

Table 6. French Nephrops trawlers, Celtic Sea (FU20-22). Discard rate (\%) by year.

| year | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| total | 65.0 | 83.8 | 58.6 | 51.2 | 86.2 | 82.0 | 60.9 | 55.8 | 63.4 | 54.3 | 65.4 | 40.1 | 40.3 | 31.7 | 64.9 | 37.4 | 49.3 | 40.7 | 28.8 | 28.7 |
| males | 46.5 | 67.0 | 38.5 | 32.8 | 73.7 | 65.3 | 40.7 | 37.0 | 44.2 | 33.6 | 45.6 | 23.0 | 23.8 | 19.8 | 46.4 | 21.0 | 30.0 | 24.0 | 16.6 | 18.2 |
| females | 86.7 | 96.5 | 86.1 | 79.6 | 96.0 | 96.3 | 90.2 | 82.3 | 88.3 | 88.1 | 94.7 | 75.0 | 72.9 | 55.6 | 85.5 | 80.8 | 90.6 | 81.4 | 68.8 | 48.9 |

As presented above, the model based on probability density with skewness gives generally adequate results (see parameters' CV) except for three years on twenty of the overall time-series. Nevertheless, the provided CV are estimated by the model and do not necessarily reflect the actual uncertainty because of complex organisation of samples (sub-sampling stratified plan applied onboard). This is illustrated by the sampled year 1997 which showed high spatial and temporal variability of discard size-composition vs. that of landings (CV of samples $>30 \%$ ) although the estimated by the model CV seems unlikely (weak value of $6.21 \%$ ). Moreover, the generated by the model total number of discarded Nephrops for 1997 was under-estimated ( 66 millions i.e. $68 \%$ of the total number estimated by sampling: 97 millions). The use of the coefficient $\gamma$ in the model was justified by the expected skewness of discard distributions due to the selectivity effect: in fact, all values of $\gamma$ do not exceed 1 . However, using the simulated model for the year 1997 with assumed symmetrical distribution of dis-
cards and with no constraint on relationship between mean sizes in discards and in landings provided more satisfactory results (Figure 17). The symmetrical simulation gave un estimate of 83 millions of discards i.e. $86 \%$ of the 97 millions calculated by sampling closer than the value generated with skewness. Moreover, the CV of parameters $\alpha, \mathrm{Lm}$ and mainly $\beta$ are less strong.

There is no current statistical evidence for choosing symmetrical or not distribution for simulations and there is no possibility to validate any relationship between mean sizes in discards and landings while the actual sampling is limited to only one complete year.

However, as underlined in the Stock Annex, the generated by model cpue (including discards calculated by the probabilistic simulation with skewness) show a good agreement with EVHOE groundfish survey indices for the period 1997-2005 ( $\mathrm{R}^{2}=0.65$ ) whilst the relationship between lpue and EVHOE indices seems more sparse ( $\mathrm{R}^{2}=0.36$ ). As also reported by WGSSDS 2007, throughout the overall time-series, some high (years 1988, 2001) or low (year 1990) values of simulated discard rates coincide with increase or decrease of lpue for 1-2 years later (increase in 1989-1990 and 2002-2003, decrease in 1991-1992). It is noticeable that no constraint was set for back-calculations on the relationship between discard rate (year i) and lpue (years $\mathrm{i}+1 / \mathrm{i}+2$ ).

## Estimates of Irish discards

Estimates of Irish discards by quarter (since 2002), total numbers of discarded individuals, parameters $\alpha, \beta$ and Lm and corresponding coefficients of variation (CV, in $\%$ ), are provided below (Table 7).

A first examination of results shows an overall better statistical adequacy than for French discards. Except for one sampled quarter (coloured fonts; 2005-Q2), the coefficients of determination are strong and the CV of model parameters remain relatively low. Despite this initial overview, the adequacy of the probabilistic approach will be tested as regards the procedure developed for the Bay of Biscay stock.

The Table 8 and Figure 18 present quarterly discard rates by sex and combined for the overall time-series. Discard rates by sampling and by simulation can be directly compared for 11 quarters (Table 8): it seems that the average simulated discard percentage is slightly lower than the sampled one ( $26.0 \%$ against $27.3 \%$ ), but for 8 quarters on 11, the simulated values are under-estimated.

The Table 9 and Figure 19 give comparisons between sampled and simulated discarded numbers. Two sampled years (2003 and 2005) for the 1st quarter give low correlations between sampled and simulated discards. Despite more good correlation levels ( 9 on 11), the overall conclusion is that the null hypothesis (slope=1) is refused apart from one example (2004-Q2) which although provides biased results of simulated discards (very high ratio Nexp/Nobs). It is worth noting that the descending part of simulated DLF of discards seems to be more coherent with the sampled DLF than the ascending one (except for one case on $11,2005-\mathrm{Q} 2$ which is denoted by the less good statistical consistency of simulation in regards with the low value of @2: Table 7). Introduction of some constraint between mean sizes in discards and in landings as for the French example may give different results for the ascending DLF.

Table 7. Irish Nephrops trawlers, Celtic Sea (FU20-22). Estimates of discards, coefficients of model and coefficients of variation of parameters (bold characters=sampled quarters).

| year | Q | disc | Lm | CV(Lm $)$ | $\boldsymbol{\alpha}$ | $\mathrm{CV}(\boldsymbol{\alpha})$ | $\boldsymbol{\beta}$ | $\mathrm{CV}(\boldsymbol{\beta})$ | $\boldsymbol{\rho}^{\mathbf{2}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | Q1 | 2664 | 26.039 | 0.95 | 1282 | 13.89 | 0.674 | 18.09 | 0.990 |
| 2003 | Q1 | 6318 | 20.994 | 1.97 | 1476 | 11.52 | 0.319 | 15.53 | 0.855 |
| 2004 | Q1 | 2208 | 24.743 | 1.34 | 998 | 18.48 | 0.625 | 24.42 | 0.960 |
| 2005 | Q1 | 7613 | 25.929 | 0.88 | 3764 | 13.27 | 0.691 | 17.29 | 0.994 |
| 2006 | Q1 | 11279 | 25.218 | 0.68 | 4594 | 8.56 | 0.564 | 11.32 | 0.929 |
| 2002 | Q2 | 1670 | 27.891 | 1.10 | 666 | 14.69 | 0.555 | 19.37 | 0.950 |
| 2003 | Q2 | 10236 | 25.119 | 0.72 | 4204 | 8.98 | 0.571 | 11.84 | 0.980 |
| 2004 | Q2 | 4953 | 24.685 | 1.05 | 1003 | 6.39 | 0.278 | 8.59 | 0.951 |
| 2005 | Q2 | 23437 | 25.139 | 1.42 | 3701 | 6.79 | 0.214 | 9.27 | 0.608 |
| 2006 | Q2 | 15977 | 26.854 | 0.35 | 7902 | 5.61 | 0.688 | 7.35 | 0.987 |
| 2002 | Q3 | 729 | 27.444 | 0.77 | 363 | 13.40 | 0.686 | 17.73 | 0.982 |
| 2003 | Q3 | 15985 | 22.042 | 0.43 | 5780 | 4.04 | 0.504 | 5.33 | 0.940 |
| 2004 | Q3 | 1291 | 28.143 | 0.26 | 571 | 3.90 | 0.615 | 5.13 | 0.969 |
| 2005 | Q3 | 4795 | 24.751 | 0.64 | 2562 | 10.55 | 0.739 | 13.85 | 0.960 |
| 2006 | Q3 | 2518 | 25.484 | 0.44 | 1144 | 6.48 | 0.626 | 8.60 | 0.927 |
| 2002 | Q4 | 11343 | 24.442 | 0.56 | 5197 | 7.89 | 0.631 | 10.46 | 0.990 |
| 2003 | Q4 | 2166 | 24.284 | 0.83 | 630 | 7.23 | 0.402 | 9.64 | 0.967 |
| 2004 | Q4 | 1561 | 27.543 | 0.93 | 713 | 14.91 | 0.630 | 19.77 | 0.992 |
| 2005 | Q4 | 9249 | 24.318 | 0.67 | 4603 | 10.22 | 0.687 | 13.49 | 0.992 |
| 2006 | Q4 | 10394 | 25.289 | 0.67 | 5666 | 11.50 | 0.753 | 15.11 | 0.990 |

Table 8. Irish Nephrops trawlers, Celtic Sea (FU20-22). Discard rate (\%) by quarter and year (for the sampled quarters: the cited percentages in bold correspond to the sampling results; those in brackets are obtained by the simulation).

| year | 2002 | 2003 | 2004 | 2005 | 2006 | 2002 | 2003 | 2004 | 2005 | 2006 | 2002 | 2003 | 2004 | 2005 | 2006 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| quarter | Q1 | Q1 | Q1 | Q1 | Q1 | Q2 | Q2 | Q2 | Q2 | Q2 | Q3 | Q3 | Q3 | Q3 | Q3 | Q4 | Q4 | Q4 | Q4 | Q4 |
| total | 7.3 | 26.9 | 15.4 | 35.3 | 41.1 | 2.6 | 37.6 | 11.5 | 21.4 | 29.5 | 1.2 | 41.2 | 10.1 | 11.1 | 19.5 | 9.9 | 26.4 | 2.3 | 54.3 | 7.2 |
|  |  | (41.6) |  | (24.5) | (32.4) |  | (29.9) | (16.5) | (28.8) | (24.1) |  | (40.6) | (9.0) |  | (15.6) |  | (22.9) |  |  |  |
| males | 6.6 | 22.1 | 13.7 | 37.9 | 34.5 | 2.5 | 34.0 | 11.1 | 19.3 | 22.9 | 1.3 | 42.2 | 9.3 | 5.2 | 17.0 | 10.9 | 20.7 | 4.3 | 47.0 | 8.0 |
| females | 8.9 | 75.1 | 18.7 | 34.0 | 56.8 | 2.7 | 40.5 | 11.7 | 22.7 | 32.7 | 1.2 | 40.6 | 11.4 | 40.0 | 20.9 | 6.5 | 59.1 | 0.2 | 71.2 | 3.8 |

It would also be interesting to re-examine the comparisons after assuming skewness of discards distributions (use of coefficient $\gamma \neq 1$ as for the French fleet). It is noticeable that for 5 quarters on 11 (Figure 19) the DLF of samples deviates from the assumed symmetry of simulations, then small sized individuals are under-estimated (however, the overestimation of the small Nephrops by the simulation occurs less often, but provides extremely divergent results). Although, there is no current basis for further analysis of this point because there is no evidence of any particular effect of some biological feature affecting the symmetry of distributions i.e. moulting which occurs in spring and autumn (example examined in the French fishery of the Bay of Biscay). The short time-series and the low sampling rate do not allow generalising this first overview.

Table 9. Irish Nephrops trawlers, Celtic Sea (FU20-22). Relationships between discarded numbers by sampling (Nobs) and by simulation (Nexp).

| year/quarter | Nexp $=\boldsymbol{\Psi}$ (Nobs) | $\boldsymbol{p}^{\mathbf{2}}$ | p(slope) | Nexp/Nobs |  |
| :---: | :---: | :--- | :---: | :---: | :---: |
| 2003 | Q1 | Nexp $=0.87^{*}$ Nobs +84.99 | 0.44 | 0.41 | $194 \%$ |
| 2005 | Q1 | Nexp $=0.60^{*}$ Nobs-2.72 | 0.72 | $0.00^{*}$ | $60 \%$ |
| 2006 | Q1 | Nexp $=0.72^{*}$ Nobs- 12.49 | 0.89 | $0.00^{*}$ | $69 \%$ |
| 2003 | Q2 | Nexp $=0.72^{*}$ Nobs-3.87 | 0.84 | $0.00^{*}$ | $71 \%$ |
| 2004 | Q2 | Nexp $=0.94^{*}$ Nobs +45.90 | 0.85 | 0.38 | $152 \%$ |
| 2005 | Q2 | Nexp $=0.78^{*}$ Nobs +267.45 | 0.85 | $0.00^{*}$ | $148 \%$ |
| 2006 | Q2 | Nexp $=0.83^{*}$ Nobs-39.77 | 0.94 | $0.00^{*}$ | $76 \%$ |
| 2003 | Q3 | Nexp $=0.89^{*}$ Nobs +32.24 | 0.94 | $0.00^{*}$ | $97 \%$ |
| 2004 | Q3 | Nexp $=0.86^{*}$ Nobs +0.92 | 0.97 | $0.00^{*}$ | $88 \%$ |
| 2006 | Q3 | Nexp $=0.80^{*}$ Nobs-2.90 | 0.91 | $0.00^{*}$ | $77 \%$ |
| 2003 | Q4 | Nexp $=0.74^{*}$ Nobs +5.79 | 0.88 | $0.00^{*}$ | $83 \%$ |

Note: *=significant result (1- $\alpha=0.95$ ).

## Conclusion

The biolognical sampling onboard for Nephrops FU20-22 stock remains poor for both main fleets. The duration of trips for French trawlers (12-15 days) restricts possibilities of regular participation of observers. Moreover, in agreement with results of sampling design applied in 1997, the long duration of trips implies a high spatial variability of harvested areas by trip and a low total number of trips sampled by quarter. Thus, the CV of discarded numbers estimated by sampling remains high. By the way, the simulations developed on French discards are hampered by the sampling of only one year throughout a long time-series. The discard practices during the whole period may change, but there is no current possibility to test the effect of such a modification on the hand-sorting onboard. In spite of that, some discard rates by year agree overall with independent indices as EVHOE groundfish survey indices (as pointed by last year's WG) and with the most notable changes in terms of lpue during the whole time-series.

The Irish dataset takes more promising because of a shorter duration of trips. Hence, conceptual problems of sampling design inherent to the French fleet should not affect the Irish data. As the Irish fleet seems to be more recruitment directed, the indices provided by the sampling onboard should improve the diagnostic accuracy. In the meantime, the simulation based on the probabilistic approach indicated an overall consistent reconstitution of discards for more sampled quarters. Many further investigations have to be carried out in the order to validate extrapolations from French catches to Irish for the period before 2002.

## B.3. Surveys

Direct Nephrops assessment by trawling is inappropriate because of notable diurnal variations of availability which is higher during dawn and dusk. The most adapted way is based on transect with video and TV runs of burrows (combined with hauls on area and geo-statistical analysis of catches with the aim of separating burrows of Nephrops from those of squat lobster), but it needs heavy preliminary arrangements because the spatial heterogeneity of resource requires to well define the survey area and the sampling plan in order to avoid biased results. The current situation will be improved in the future once a data time-series has been collected by the Irish specifically designed survey program launched in 2006. However, the Irish and French ex-
ploited areas are different. On FU20-22 the French groundfish survey EVHOE while not focusing on Nephrops does provide an indication of the length distributions and the strength of recruitment (Figure 20). An Irish groundfish survey giving size composition of Nephrops catches has also been carried out since 2003. Moreover, a UK bottom trawl survey had occurred on the same area between 1984 and 2004, but only two sampling stations were within FU20-22 area.

A comparative analysis conducted between lpue and cpue of French and Irish vessels with EVHOE indices shows a good agreement between commercial French cpue and EVHOE series for the period 1997-2005 ( $\mathrm{R}^{2}=0.65$ ) whilst the relationship is more sparse ( $\mathrm{R}^{2}=0.36$ ) when the commercial French lpue are used (Figure 21). The Irish data are not significantly linked to the French dataset probably due to the difference of harvested area and the short time-series

The results of the UWTV survey initiated by Republic of Ireland in 2006 involving in the three first years, 2006-2008, are shown by Figures 20-25 and Tables 10-11. It is noticeable that the strongest values of this short time-series (2006) coincide with the highest level on "Smalls" as reported by Irish industry in 2007. In a timeframe of around $2-4$ years, this survey should provide valuable information to tune data for the FU20-22 Nephrops stock especially on the "Smalls" ground where are located more than the $2 / 3$ of the total Irish yearly production. Nevertheless, the historical longer series of French landings in the Celtic Sea is less involved by the area covered by UWTV (the contribution of the rectangle 31E3 in the total French production fell from $41 \%$ in 1999 at less than $10 \%$ in 2008). This implies the necessity to tune data for the whole area.

## B.4. Commercial cpue

Between 2006 and 2007, the French fishing effort declined notably by $-25 \%$ and the lpue increased $(+13 \%)$ although the evolution of the same indicators for the Irish fleet was different ( $+31 \%$ of fishing effort and $+36 \%$ of lpue). It is noticeable that the decrease of the French fishing effort was caused by the reduction of the number of trips by vessel whereas the total number of vessels remained almost stable. The evolution of the Irish fishing effort involves either in increase of the fishing vessels (95 Irish trawlers were listed in 2007 against 80 for 2006) or in increase of the number of trips by vessel.

Between 2007 and 2008, the effort of the French trawlers decreased slightly i.e. 99789 h against 101980 h for 2007 whereas the Irish fishing effort remained stable ( 59727 h against 59899 h in 2007). Lpue of both fleets increased mainly for French trawlers $(+22 \%: 22.6 \mathrm{~kg} / \mathrm{h}$ against $18.5 \mathrm{~kg} / \mathrm{h}$ for 2007 ) and, to a lesser degree, for Irish $(+11 \%$ : $55.2 \mathrm{~kg} / \mathrm{h}$ against 49.4 in 2007).

## C. Historical stock development

There is no currently specific development for analytical assessment of the stock. By the WGNEPH 2003, the FU20-22 Nephrops stock was analytically assessed by XSA (software VPA; Darby and Flatman, 1994). Because of the lack of long and consistent Irish series (before DCR), the analysis was limited on the male component involved by French trawlers (see input parameters: Table 1)

## D. Short-term projection

No short-term projection is performed for this stock.

## E. Medium-term projections

No medium-term projection is performed for this stock.

## F. Long-term projections

No long-term projection is performed for this stock.

## G. Biological reference points

There is no biological reference point for this stock.

## H. Other issues

None.

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Figure 1. Functional units 20-22 (Nephrops grounds in Celtic Sea).


Figure 2. Nephrops FU20-22 (Celtic Sea). Spatial distribution of landings of the main fleets (average value of the period 1996-1999).



Figure 3. Nephrops FU20-22 (Celtic Sea). Above: Spatial and by year distribution of Irish landings. Below: Contribution of the rectangle 31E3 (concentrating more than $2 / 3$ of the total Irish production) in the total French landings. Years 1999-2008.











Figure 4. Nephrops FU20-22 (Celtic Sea). Spatial and monthly distribution of French landings.


Figure 5. Nephrops FU20-22 (Celtic Sea). Spatial distribution of French landings in 2007.


Figure 6. Nephrops FU20-22 (Celtic Sea). Number of Irish trawlers involving Nephrops landings.


Figure 7. Nephrops FU20-22 (Celtic Sea). Tailed proportion (in converted weight) in landings by month (left) and by year (right).


Figure 8. Nephrops of the Celtic Sea (VIIfgh, FU20-22). Biometric relationships (CL vs. 2nd abdominal segment by sex). Data harvested during the survey EVHOE 2007.


Figure 9. Nephrops of the Celtic Sea (VIIfgh, FU20-22). French landings for 2008. Length distributions (1) including the data on tails and (2) using the previous method (no sampling of tails; the total tailed proportion was apportioned in the smallest category of entire Nephrops at auction).


Figure 10. Nephrops of FU23-24 (Bay of Biscay). Final results of logistic derivation of discards. Relationship between mean sizes of landings and discards. The triangular fonts represent the results of the status quo (proportional derivation) method. The underlined years correspond to the available datasets of sampling onboard. The rhombus fonts correspond to the logistic derivation. The dark curve is provided by the final fitting on the whole time-series. The bright curve is the result of the fitting on the years with available data.


Figure 11. Nephrops of FU23-24 (Bay of Biscay). Comparison between discard rates obtained by previous (proportional) derivation and by logistic derivation. Combined sexes and whole year datasets.


Figure 12. Nephrops of FU23-24 (Bay of Biscay). Comparison between distributions of length frequencies (carapace length, CL in mm ) of discards obtained by sampling and by simulation (broken lines).


Figure 13. Nephrops of FU23-24 (Bay of Biscay). Comparison between discarded numbers of individuals obtained by simulation ( Y axis) and by sampling ( X axis). Statistical tests on linear regressions of Y vs. X by year.


| year | Da | Dobs | \% |
| :---: | ---: | ---: | ---: |
| 2005 | 0.113 | 0.101 | 85 |
| 2004 | 0.127 | 0.048 | 107 |
| 2003 | 0.135 | 0.031 | 100 |
| 1998 | 0.154 | 0.049 | 106 |
| 1991 | 0.157 | 0.044 | 97 |
| 1987 | 0.115 | 0.052 | 108 |

Figure 14. Nephrops of FU23-24 (Bay of Biscay). Statistical test (Kolmogorov-Smirnov) between cumulated frequencies of sampled and simulated discards by year.


Figure 15. Nephrops FU20-22 (Celtic Sea). Different hand-sorting logistic curves by quarter, country and dataset. In 2005 no sample was collected in France during the 1st quarter and 2nd quarter providing inconsistent results.


Figure 16. Nephrops of FU20-22 (Celtic Sea). Comparison between discard rates obtained by previous (proportional) derivation (used by WGNEPH until 2004) and by logistic derivation. Combined sexes and whole year datasets.

Nexp $=0.84^{\star}$ Nobs $+54.76 \rho^{2}=0.85 p($ slope $)=0.01[86 \%]$


| year | disc | Lm | CV $(\operatorname{Lm})$ | $\boldsymbol{\alpha}$ | CV $(\alpha)$ | $\beta$ | CV $(\beta)$ | $\rho^{2}$ |
| :---: | ---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 83306 | 29.807 | 1.29 | 32335 | 9.42 | 0.538 | 6.43 | 0.913 |

Figure 17. Nephrops of FU20-22 (Celtic Sea). French fleet. Results of the discard simulation on the year 1997. The distribution is assumed symmetrical and no constraint was set on relationship between mean sizes in discards and landings. Simulated number (Nexp) illustrated by broken line are compared to sampled one (Nobs).


Figure 18. Nephrops of FU20-22 (Celtic Sea). Discard rate (\%) of Irish trawlers by year and quarter.


Figure 19. Nephrops FU20-22 (Celtic Sea). Irish trawlers . DLF of sampled (continuous line) and simulated (broken line) discarded numbers.


Figure 20. Nephrops FU20-22. Indices of the French groundfish survey EVHOE.


Figure 21. Nephrops FU20-22. Comparison of indices EVHOE and of commercial lpue and cpue for French and Irish trawlers.


Figure 22. Omnidriectional mean variograms for the Celtic Sea FU20-22 by year from 2006-2008.


Figure 23. Cross validation plots for the Celtic Sea FU20-22 by year from 2006-2008.


Figure 24. Contour plots of the krigged density estimates for the Celtic Sea FU20-22 by year from 2006-2008.


## FU20-22 Burrow density (no/m2)

Figure 25. Burrow density distributions for the Celtic Sea FU20-22 by year from 2006-2008.

Table 10. Summary geostatistics for the Nephrops UWTV surveys of the Celtic Sea from 20062008.

| Ground | Year | Number <br> of stations | Number <br> of <br> boundary <br> points | Mean <br> Density <br> (No./M2) | Standard <br> Deviation | CVgeo <br> $(\%)$ | Var | Domain Area (m2) | Raised abundance <br> estimate (million <br> burrows) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Smalls | 2006 | 100 | 50 | 0.62 | 0.50 | $80 \%$ | 0.25 | 2847 | 1914 |
| Smalls | 2007 | 107 | 63 | 0.46 | 0.44 | $96 \%$ | 0.19 | 2915 | 1402 |
| Smalls | 2008 | 76 | 31 | 0.47 | 0.40 | $85 \%$ | 0.16 | 2698 | 1448 |

Table 11. Summary statistics for the Nephrops UWTV survey indicator stations of the Labadie and Nymphe Bank and Seven Heads Grounds from 2006-2008.

| Ground | Year | Number <br> of stations | Mean <br> Density <br> $($ No./M2) | Area <br> Surveyed <br> $(\mathrm{M} 2)$ | Burrow <br> count | Standard <br> Deviation | 95\%CI | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2006 | 9 | 0.42 | 1,322 | 760 | 0.37 | 0.28 | $29 \%$ |
| Labadie Bank | 2007 | - | - | - | - | - | - | - |
|  | 2008 | - | - | - | - | - | - | - |
| Nymphe Bank | 2006 | 2 | 0.27 | 195 | 89 | 0.39 | 3.47 | $100 \%$ |
|  | 2007 | - | - | - | - | - | - | - |
| Seven Heads | 2008 | - | - | - | - | - | - | - |
|  | 2006 | 7 | 0.23 | 995 | 293 | 0.25 | 0.23 | $41 \%$ |
|  | 2008 | - | - | - | - | - | - | - |

*random stratified estimates are given for the Labadie Bank, Nymphe Bank and Seven Heads grot

- Area not surveyed in 2007 to 2008 due to weather


## Stock Annex 7.10: Plaice in VIIfg

- Stock Annex 7.10 Plaice VIIfg for latest update see WGCSE 2009, Annex 07.10 Plaice VIIfg


## Stock Annex 7.13: Sole in VIIfg

- Stock Annex 7.13 Sole VIIfg for latest update see WGCSE 2009, Annex 07.13 Sole VIIfg


## Stock Annex 7.15: Whiting VIIe-k

Stock specific documentation of standard assessment procedures used by ICES.

| Stock | Whiting VIIe-k |
| :--- | :--- |
| Working Group | Celtic Sea Eco-region |
| Date | 17 May 2010 |
| Revised by | Sarah Davie |

## A. General

## A.1. Stock definition

The degree of separation of whiting stocks between the Irish Sea, and ICES Divisions VIIb-c from the Celtic Sea, is currently unclear. SAMFISH (EU Study Contract 99-009, Improving sampling of western and southern European Atlantic Fisheries) described the stock unit as follows:

The main spawning areas of whiting in the Western Channel and Celtic Sea are off Start Point, off Trevose Head and southeast of Ireland. The spawning season is from February to May, and the larvae are found in mid-water before moving to live near the seabed by September. For the next two years, juvenile whiting are found in shallow coastal and estuarine areas, being particularly abundant around Start Point. Nearly 4000 adult whiting were tagged and released off Start Point during August 1958 and 1960. Most returns were within three months of release and demonstrated little indication of movement. Subsequent recaptures indicated more movement of whiting into the Celtic Sea than between the western and eastern Channel. Whiting released in summer between 1957 and 1961 near Carmarthen Bay moved south and west towards the two spawning grounds off Trevose and southeast of Ireland. There was no evidence of emigration out of the Celtic Sea area. Returns of whiting tagged and released in the County Down spawning area in the Irish Sea demonstrate more movement south into the Celtic Sea than north to the west of Scotland.

## A.2. Fishery

Whiting in Divisions VIIe- k are taken as a component of catches in mixed trawl fisheries. Whiting landings through the mid 1980s totalled between 10000 t and 15000 t , through the mid to late 1990s landings were elevated to around 20000 t . Since the turn of the century, landings have been in decline and are now below 10000 t . Through the 1980s and early 1990s France accounted for around $60-85 \%$ of landings. While Ireland accounted for between $10 \%$ and $20 \%$ of landings, the UK $10 \%$, and Belgium had minimal contribution ( $1-2 \%$ ). Landings from both the UK and Belgium have remained at similar levels over time. Since the early 1990s Ireland has accounted for a greater proportion of landings. Proportions since 2004 have been similar to France whose landings have been falling since the turn of the century.

French landings are made mainly by gadoid trawlers, which prior to 1980 were mainly fishing for hake in the Celtic Sea. Irish demersal trawlers from Dunmore East and Castle-
townbere and other ports in southwest Ireland have traditionally targeted Celtic Sea whiting in a mixed trawl fishery. In response to poor catches in other areas vessels have been attracted into this fishery in recent years from County Donegal.

A detailed description of the Irish fishery is given in the annual WD to WGSSDS: 'A summary of the Irish Fishery and Sampling of Whiting in VIIe-k'.

## A.3. Ecosystem aspects

No relevant information has been made available to the Working Group.

## B. Data

## B.1. Commercial catch

Data on international landings-at-age and mean weight-at-age are available for Irish, French and UK fleets from 1999 to present. The following procedures have been applied to aggregate the data for the areas VIIe, VIIfgh and VIIj,k and build the database for VIIek . UK VIIe-k data were used to scale catch numbers according to the landings for each area. French VIIf,g,h data were used with Irish VIIg data to scale VIIf,g,h catch numbers. Irish VIIj data were used to scale VIIj,k catch numbers. The Table below demonstrates the data available and the procedures used to derive quarterly length compositions, age compositions and mean weights-at-age.


## B.2. Biological

Age group 0 is included in the assessment data to allow inclusion of 0 -group indices in the XSA, although in most years, no landings are recorded. Very small landings of 0 group whiting were not included in the catch-at-age datafile to avoid spurious Fshrinkage effects at this age. Mean weights-at-age in the catch were derived by combining French, Irish and English data, weighted by the numbers landed at-age.

Mean weight-at-age in the stock are taken as mean weights-at-age in the quarter 1 catch. Where age 1 was poorly represented in quarter 1 landings, quarter 2 values were used as estimates of mean weight-at-age 1 in the stock. Stock weights-at-age are smoothed using
a three year rolling average across ages to dampen the noise exhibited by the stock weight dataset. This approach is also used in Irish Sea whiting and Celtic Sea haddock.

Natural mortality is assumed to be 0.2 over all age groups and years.
Maturity data collected in the Celtic Sea in November 2002 during the French EVHOE survey were presented to the WG (Working Document 1: WGSSDS 2003). Results indicated $13 \%$ of age 1 fish are mature, $97 \%$ at-age 2, and $100 \%$ at-age 3 and older. These results are similar to previous assumptions of knife-edged maturity at-age 2. Exploratory analyses indicated that use of the French maturity ogive made little impact on the assessment. The WG therefore retained the assumptions of knife-edged maturity at-age 2. Since 2006 the knife edge maturity ogive has been replaced with indices calculated based on data from the UK WCGFS (Working Document 3: WGSSDS 2006) but a fixed vector is still used. Maturity sampling by Ireland and the UK on dedicated surveys confirms the use of this ogive but is insufficient to provide annual data.

The proportions of F and M before spawning were both set to zero to reflect the SSB calculation date of 1 January.

The knife edge maturity ogive was replaced with new indices calculated based on data from the UK WCGFS as detailed in WD 3, WGSSDS, 2006.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5 +}$ |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Maturity | 0 | 0.39 | 0.90 | 0.99 | 0.99 | 1.00 |

## B.3. Surveys

The following surveys are available as survey tuning data input for the assessment of whiting VIIe-k:

- UK-WCGFS, 1987-2004

The March UK groundfish survey was extended in 1992 to provide better coverage for gadoids in VIIf,g. The whiting tuning data calculated from this survey is for VIIf,g. The survey was carried out on the RV Cirolana until 2003. In 2004 it was carried out on the RV Endeavour and discontinued thereafter. The survey fished fixed station positions allocated by area and depth strata. The survey used a modified Portuguese High-Headline trawl (PHHT) with 350 mm rubber bobbins, a bunt tickler chain and a 20 mm codend liner. The mean log standardized index by year demonstrated some evidence of positive catchability in the last three years of the survey (2002-2004) and cohort tracking in the mean standardized index up to then was very noisy in the last three years. These years were not included in the final assessment.

- UK-BCCSBTS-S, 1988-2001

The Autumn UK Bristol Channel beam trawl survey (VIIf) is commercially rigged (1989 style) with 4 m beam trawl fitted with a chain mat, flip-up ropes, and a 40 mm codend liner. The gear is towed at 4 knots (ground speed) for 30 minutes. This survey provides information for age 0 and age 1 whiting.

- FR-EVHOE, 1997-present

This fourth-quarter annual groundfish is carried out on the RV Thalassa. Age data are available from 2001 onwards. The sampling design is a stratified random allocation. The number of hauls per stratum is optimized by a Neyman allocation taking into account the most important commercial species in the area (hake, monkfish and megrim). The fishing gear used is a GOV with an average vertical opening of 4 m and a horizontal opening of 20 m .

- IR-WCGFS, 1993-2002

The fourth-quarter Irish west-coast groundfish survey (WCGFS) was carried out in VIaS and VIIbj on chartered commercial vessels. The sampling design attempted to allocate at least two stations per rectangle. Stations were selected randomly within each rectangle from known clear tow positions. A Rockhopper GOV with 12 inch discs was used. The nets were fitted with a 20 mm codend liner. This survey was discontinued after the 2002 survey, giving way to a new Irish groundfish survey on board the RV Celtic Explorer.

- IR-ISCSGFS, 1997-2002

Ireland commenced a Celtic Sea research vessel survey on board the RV Celtic Voyager in 1997 carried out in VIIa and VIIg. The survey used a GOV Trawl with a mean vertical opening is 6 m and door spread 48 m . Data from this survey (IR-ISCSGFS) were presented for the first time to the 2003 WG . The data made available were from prime stations only in a limited area of Division VIIg. The survey was discontinued after the 2002 survey, giving way to a new Irish groundfish survey on board the RV Celtic Explorer.

- IR-GFS 7g and j, 2003-present

Ireland commenced a new fourth quarter survey in 2003 on board the RV Celtic Explorer which covers VIaS, VIIbgj as part of the internationally coordinated, Quarter 4 IBTS survey program. The IGFS has a random stratified design and uses a GOV (with rock-hopper in VIa) with a 20 mm codend liner. This is a substantially different design to the Irish Sea/Celtic Sea groundfish survey (IR-ISCSGFS) it replaces. Data from this survey (IR-GFS) were presented for the first time to the 2004 WG .

- IR-IGFS Swept Area, 1999-present

This survey index constitutes a combination of the IR-ISCSGFS and IR-GFS surveys in the area of overlap between them (VIIg). The two surveys were standardized using a swept-area estimate of catches, described in WD 5 (WGSSDS 2006). This survey was presented for the first time to the 2006 WG . The mean standardized index by year demonstrated good tracking of the strong 1999 year class to age 7 with the exception of age 4 in 2003. Although the source data were checked, this is probably an anomaly of the year effect in 2003. This point has been removed from recent assessments to ensure the survey gets higher scaled weight in further runs. This compromise is not ideal but given the short time-series of the survey and apparently good performance otherwise the WG considered that the survey should be a good index for this stock.

## B.4. Commercial cpue

Information on effort, and whiting landings and lpue are available from a number of commercial fleets. This includes two French (gadoid and Nephrops directed) since 1983, four Irish (VIIj, and VIIg otter trawlers, and Scottish seines) since 1995, in addition to effort only from UK England and Wales VIIe-k beam trawlers and VIIe-k otter trawlers since 1983.

Across the majority of commercial fleets lpue has fallen over time, as is the case with landings. In the mid 1990s at the start of the Irish Scottish seine dataseries lpue was high, falling steeply over several years. Lpue continues to remain at these lower levels with some annual fluctuation. In relation to otter trawlers, the French gadoid directed fleet consistently revealed the highest lpue. This too has declined over the period of data available to levels half those of the early 1980s. The Irish VIIg otter trawl fleet is the only one to demonstrate an overall increasing lpue trend although the increase has been relatively small.

## B.5. Other relevant data

No other relevant data to report.

## C. Historical stock development

Data screening: Exploratory data analysis carried out using FLR. A separable VPA was performed using the Lowestoft VPA95 software to screen for outliers in the catch numbers.

Model used: XSA
Software used: FLR under R version 2.4.1 in conjunction with FLCore 1.4-3, FLAssess 1.4.1, FLXSA 1.4-2 and FLEDA 1.4-2

Lowestoft VPA95 software also for XSA and separable VPA
Model Options:

| Option | Setting |  |
| :--- | :---: | :---: |
| Ages catch dep stock size | None |  |
| Q plateau |  |  |
| Taper | No |  |
| F shrinkage SE | 1.00 |  |
| F shrinkage year range | 5 |  |
| F shrinkage age range |  | 3 |
| Fleet SE threshold |  | 0.50 |
| Prior weights | No |  |

Input data types and characteristics:

| Type | Name | Year range | Age <br> range | Variable year <br> to year |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes | 1982 -current | $0-7+$ | Yes |
| Canum | Catch-at-age in numbers | 1982 -current | $0-7+$ | Yes |
| Weca | Weight-at-age in the commercial catch | 1982 -current | $0-7+$ | Yes |
| West | Weight-at-age of the stock at spawning time | 1982 -current | $0-7+$ | Yes: |
| Mprop | Proportion of natural mortality before spawning | 1982 -current | $0-7+$ | No |
| Fprop | Proportion of fishing mortality before spawning | 1982 -current | $0-7+$ | No |
| Matprop | Proportion mature-at-age | 1982 -current | $0-7+$ | No |
| Natmor | Natural mortality | 1982 -current | $0-7+$ | No |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :--- | :---: |
| Tuning fleet 1 | FR-Gadoid Late | 1993 -current | $3-6$ |
| Tuning fleet 2 | FR-Nephrops | 1993 -current | $3-6$ |
| Tuning fleet 3 | FR-EVHOE | 1997 -current | $0-4$ |
| Tuning fleet 4 | UK-WCGFS | 1987 -current | $1-6$ |
| Tuning fleet 5 | IR-IGFS Swept area | 1999-current | $0-6$ |

Settings for each assessment since 1999 are detailed in Table 1. Trial runs have, over the years, explored most of the options with regards XSA settings. This stock has not had a benchmark assessment, however exploratory assessments have been carried out within the WGSSDS up until 2007.

## D. Short-term projection

Model used: Multi Fleet Deterministic Projection
Software used: MFDP1a
Initial stock size: initial stock numbers derived from XSA analyses. Numbers-atage 0 are not considered to be well estimated and are replaced with a geometric mean of the full time-series (1982-2007). Recruitment has been at a low level since 1995 with the exception of the 1999 year class. The two most recent years have displayed good recruitment, with last year's being revised downward. Recruitment is solely estimated from the FR-EVHOE and IR-GFS7gSweptArea surveys, in recent years the French survey estimates have been far higher than those of the Irish survey. Because of these reasons the geometric mean is used.

Natural mortality: That used in the assessment
Maturity: Maturity ogive used in the assessment
F and $M$ before spawning: Those used in the assessment method
Weight-at-age in the stock: Unscaled 3 year arithmetic mean

Weight-at-age in the catch: Unscaled 3 year arithmetic mean
Exploitation pattern: Unscaled 3 year arithmetic mean (though alternative options may be used depending on recent F trajectories and the Working Group's perception of the fishery).

Intermediate year assumptions: Status quo F
Stock-recruitment model used: Geometric mean of full time-series (1982 to pre-sent-1) for age 0 recruitment

Fbar: That used in the assessment

## E. Medium-term projections

None.

## F. Long-term projections

Model used: Multi Fleet Yield-per-recruit
Software used: MFYPR2a
Yield-per-recruit calculations are conducted using the same input values as those used for the short-term forecasts.

## G. Biological reference points

A summary of reference point proposals to date, their technical basis and currently adopted reference points is given in the text Table below:

|  | WG 1998 | ACFM 1998 | WG 2000 | ACFM 2000 |
| :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}_{\text {lim }}$ | No Proposal | No Proposal | $1.18{ }_{\text {(Fim=Floss) }}$ | No Proposal |
| $\mathrm{F}_{\mathrm{pa}}$ | No Proposal | No Proposal | $0.72\left(F_{p a}=F_{\text {lim }} \times \mathrm{e}^{-1.645 \times 0.3}\right)$ | No Proposal |
| $\mathbf{B}_{\text {lim }}$ | 15,000 t | 15,000 t | $15000 \mathrm{t}_{\text {(Blim=Bloss) }}$ | 15,000 t (Blim=Bloss) |
| $\mathrm{B}_{\mathrm{pa}}$ | 18,000 t | 21,000 t | $21000 \mathrm{t}_{\text {(Bpa=Bloss } \times 1.4)}$ | 21,000 ${ }_{\text {(Bpa=Bloss } \times 1.4)}$ |

The technical basis of ACFM's 1998 Bpa proposal is given below (1999 WG text):
$\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } \mathrm{x} 1.4=21000 \mathrm{t}$. In the past the WG have selected MBAL as 18000 t based on evidence of reduced recruitment at SSB's $<18000 \mathrm{t}$. However this MBAL is driven by a period of low recruitments at low SSB in the earlier years of the time-series (1982-1985) when the data are probably not reliable. Examination of the stock-recruit plot provides no compelling evidence of reduced recruitment below SSB of 18000 t .

The technical basis of the WG's 2000 Flim and $\mathrm{F}_{\text {pa }}$ proposals are given below:
On the basis of results obtained from a LOWESS fitted non-parametric stock and recruitment relationship and the derived equilibrium SSB and yield curves with the original data trajectories the 2000 Working Group considered that $F_{p a}$ and Flim could be defined because Floss appeared reasonably estimated. However, taking into account the uncertainties in the data the 2000 Working Group decided to use 0.3 as the SE in calcula-
tion of $\mathrm{F}_{\mathrm{pa}}$ from $\mathrm{F}_{\text {loss. }}$. The technical basis for the proposed reference points are defined below:
$\mathrm{F}_{\text {lim }}=\mathrm{F}_{\text {loss }}$ (1.18 in this year's assessment)
$\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim } \times \mathrm{e}-1.645^{*} 0.3=0.72$
The currently adopted reference points are as follows:

| Current Reference Points |  |
| :--- | :--- |
| $\mathbf{F}_{\lim }$ | No Proposal |
| $\mathbf{F}_{\mathrm{pa}}$ | No Proposal |
| $\mathbf{B}_{\text {lim }}$ | $15,000 \mathrm{t} \quad\left(\mathrm{B}_{\mathrm{LIM}}=\mathrm{B}_{\mathrm{Loss} 1983}, \mathrm{ACFM}_{1998}\right)$ |
| $\mathbf{B}_{\mathrm{pa}}$ | $21,000 \mathrm{t} \quad\left(\mathrm{B}_{\mathrm{PA}}=\mathrm{B}_{\mathrm{LOSS} 1983} \times 1.4\right)$ |

## H. Other issues

No other issues.

## I. References

Table 1. Model settings/Input data/Tuning data.


## Stock Annex 8.2: Plaice in VIIe

- Stock Annex 8.2 Plaice VIIe: for latest update see WGCSE 2009, Annex 08.2 Plaice VIIe


## Stock Annex 8.3: Sole in VIIe

- Stock Annex 8.3 Sole VIIe: for latest update see WGCSE 2009, Annex 08.3 Sole VIIe


[^0]:    *2010 and 2011 values are TSA-derived projections of population numbers.

[^1]:    Values are tonnes.

[^2]:    * figures are provisional.

[^3]:    * provisional na $=$ not available.

[^4]:    * provisional na = not available.

[^5]:    * provisional na = not available.

[^6]:    *harvest rates previous to 2006 are unreliable

[^7]:    * provisional na = not available.

[^8]:    * Preliminary

[^9]:    revised

[^10]:    Functional units in VIIa

[^11]:    * provisional

[^12]:    ${ }^{\text {a }}$ 1989-onwards Northern Ireland included with England and Wales.
    ${ }^{\mathrm{b}}$ Based on UK(N.Ireland) and Ireland data.

    * Preliminary.

[^13]:    * Average of 1995-99
    * Average of 1995-99
    ** Includes IRL OTB VIIgj, FR GAD VIIfgh and UK Trawl VIIe-k
    ${ }_{* * *}$ assuming average effort 2006-8 for France

[^14]:    *random stratified estimates are given for the Slyne Head and Galway Bay grounds

[^15]:    ${ }^{1}$ where Nephrops constituted $10 \%$ of the landed value.
    ${ }^{2}$ A threshold of $30 \%$ of Nephrops in reported landings by trip is used to identify the landing and effort of this fleet.

[^16]:    ${ }^{\text {a }}$ : Preliminary
    ${ }^{\mathrm{b}}$ : Preliminary, Reported as VIIb-k

[^17]:    Input units are thousands and kg - output in tonnes

[^18]:    ${ }^{1}$ Estimated by the Working Group.
    ${ }^{2}$ Divisions VIId, $=4,739 \mathrm{t}$.
    ${ }^{3}$ Included in Division VIId
    ${ }^{4}$ Migration correction (15\% of VIId Qtr 1) added to stock.

[^19]:    ${ }^{a}$ XSA estimate (5252) replaced with GM recruitment69-07

[^20]:    $a_{\text {replaced }}$ XSA estimate (5252) with GM recruitment69-07

[^21]:    ${ }^{1}$ The six trips sampled in 2005 provided new s-shaped curves of hand-sorting for Q3 and Q4 which were used for simulations of the recent period since 2000 i.e. since the mesh size change.

[^22]:    ${ }^{2}$ This procedure is performed only on Irish dataset whereas it is not pertinent for French data (only one year dataset).

[^23]:    ${ }^{3}$ For French discards, are also included in the optimisation algorithm, the parameters $\alpha$ and L50 of the first period (1987-1990) which remained unknown.
    ${ }^{4}$ DDL is equal to nc-4 for French discards, but equal to nc-3 for Irish data (parameter $\gamma$ is omitted).

